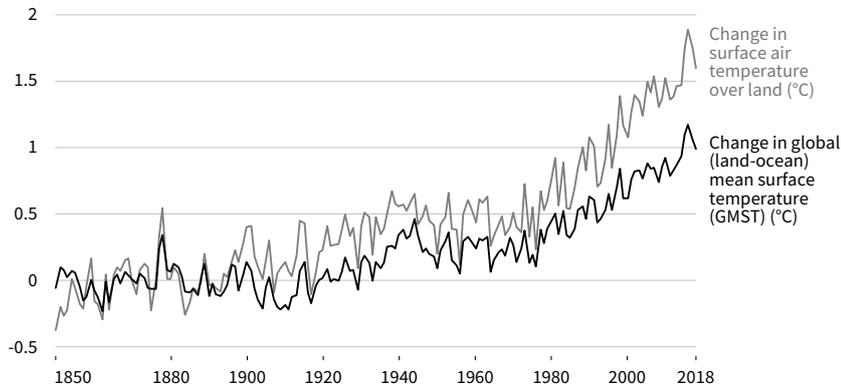


Land use and observed climate change

A. Observed temperature change relative to 1850-1900

Since the pre-industrial period (1850-1900) the observed mean land surface air temperature has risen considerably more than the global mean surface (land and ocean) temperature (GMST).

CHANGE in TEMPERATURE rel. to 1850-1900 (°C)



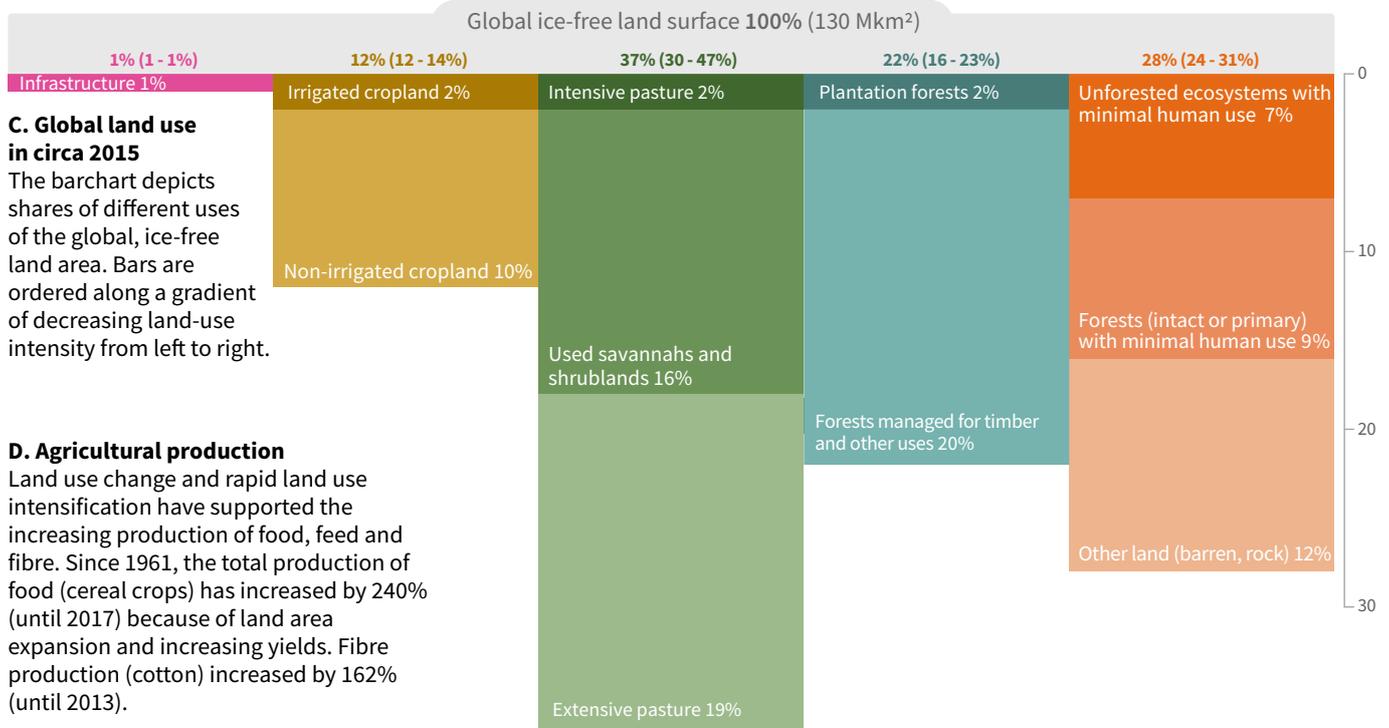
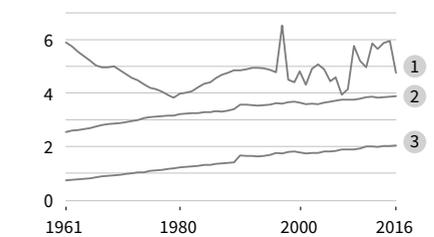
B. GHG emissions

An estimated 23% of total anthropogenic greenhouse gas emissions (2007-2016) derive from Agriculture, Forestry and Other Land Use (AFOLU).

CHANGE in emissions rel. to 1961

- ① Net CO₂ emissions from FOLU (Gt CO₂/yr)
- ② CH₄ emissions from Agriculture (Gt CO₂eq/yr)
- ③ N₂O emissions from Agriculture (Gt CO₂eq/yr)

Gt CO₂eq/yr



C. Global land use in circa 2015

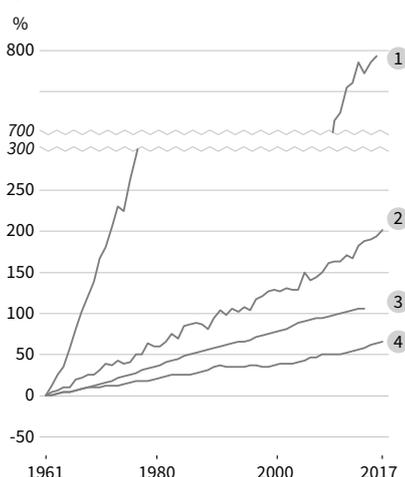
The bar chart depicts shares of different uses of the global, ice-free land area. Bars are ordered along a gradient of decreasing land-use intensity from left to right.

D. Agricultural production

Land use change and rapid land use intensification have supported the increasing production of food, feed and fibre. Since 1961, the total production of food (cereal crops) has increased by 240% (until 2017) because of land area expansion and increasing yields. Fibre production (cotton) increased by 162% (until 2013).

CHANGE in % rel. to 1961

- ① Inorganic N fertiliser use
- ② Cereal yields
- ③ Irrigation water volume
- ④ Total number of ruminant livestock

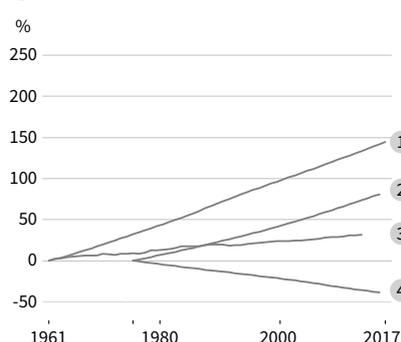


E. Food demand

Increases in production are linked to consumption changes.

CHANGE in % rel. to 1961 and 1975

- ① Population
- ② Prevalence of overweight + obese
- ③ Total calories per capita
- ④ Prevalence of underweight

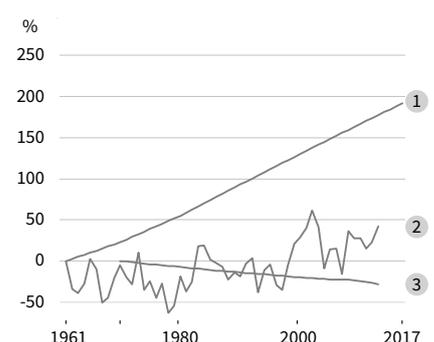


F. Desertification and land degradation

Land-use change, land-use intensification and climate change have contributed to desertification and land degradation.

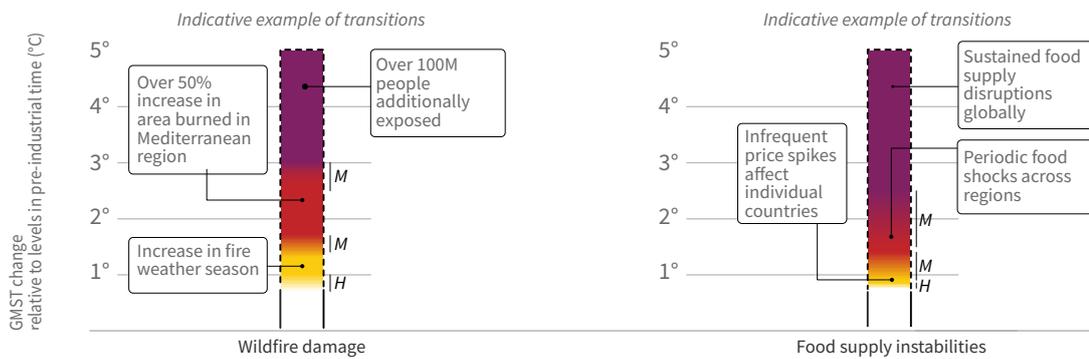
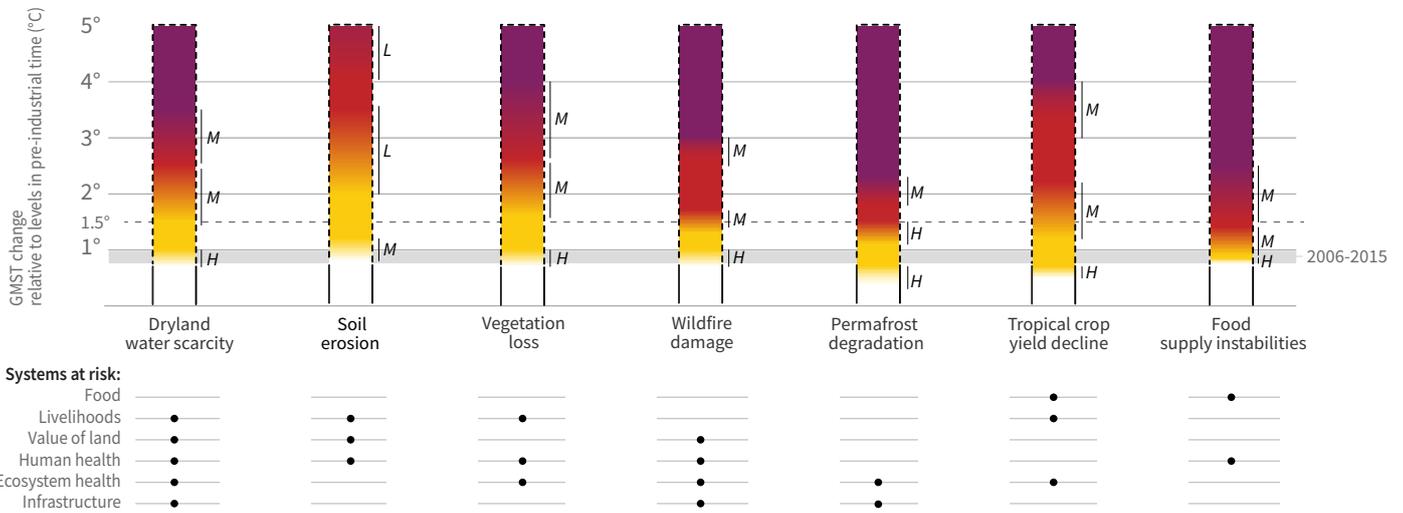
CHANGE in % rel. to 1961 and 1970

- ① Population in areas experiencing desertification
- ② Dryland areas in drought annually
- ③ Inland wetland extent

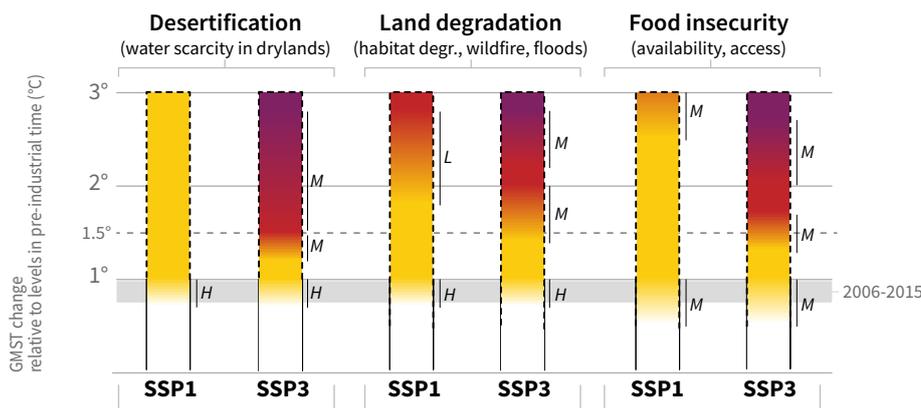


A. Risks to humans and ecosystems from changes in land-based processes as a result of climate change

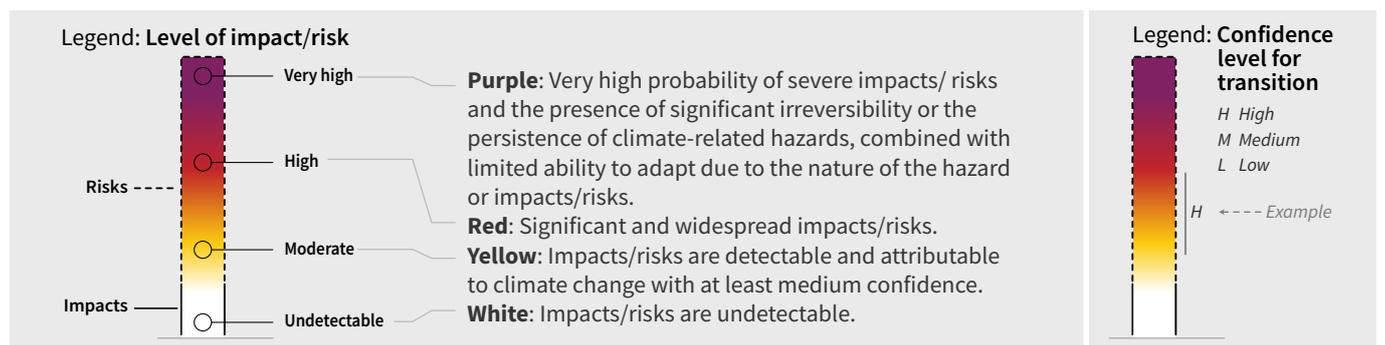
Increases in global mean surface temperature (GMST), relative to pre-industrial levels, affect processes involved in **desertification** (water scarcity), **land degradation** (soil erosion, vegetation loss, wildfire, permafrost thaw) and **food security** (crop yield and food supply instabilities). Changes in these processes drive risks to food systems, livelihoods, infrastructure, the value of land, and human and ecosystem health. Changes in one process (e.g. wildfire or water scarcity) may result in compound risks. Risks are location-specific and differ by region.



B. Different socioeconomic pathways affect levels of climate related risks



Socio-economic choices can reduce or exacerbate climate related risks as well as influence the rate of temperature increase. The **SSP1** pathway illustrates a world with low population growth, high income and reduced inequalities, food produced in low GHG emission systems, effective land use regulation and high adaptive capacity. The **SSP3** pathway has the opposite trends. Risks are lower in SSP1 compared with SSP3 given the same level of GMST increase.



Chapter 5: Food Security

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21

Coordinating Lead Authors: Cheikh Mbow (Senegal), Cynthia Rosenzweig (The United States of America)

Lead Authors: Luis G. Barioni (Brazil), Tim G. Benton (United Kingdom), Mario Herrero (Australia/Costa Rica), Murukesan Krishnapillai (Micronesia/India), Emma Liwenga (Tanzania), Prajal Pradhan (Germany/Nepal), Marta G. Rivera-Ferre (Spain), Tek Sapkota (Canada/Nepal), Francesco N. Tubiello (The United States of America/Italy), Yinlong Xu (China)

Contributing Authors: Erik Mencos Contreras (The United States of America/Mexico), Joana Portugal Pereira (United Kingdom), Julia Blanchard (Australia), Jessica Fanzo (The United States of America), Stefan Frank (Austria), Steffen Kriewald (Germany), Gary Lanigan (Ireland), Daniel López (Spain), Daniel Mason-D’Croz (The United States of America), Peter Neofotis (The United States of America), Laxmi Pant (Canada), Renato Rodrigues (Brazil), Alex C. Ruane (The United States of America), Katharina Waha (Australia)

Review Editors: Nouredine Benkeblia (Jamaica), Andrew Challinor (United Kingdom), Amanullah Khan (Pakistan), John Porter (United Kingdom)

Chapter Scientists: Erik Mencos Contreras (The United States of America/Mexico), Abdoul Aziz Diouf (Senegal)

Date of Draft: 07/08/2019

| | | |
|----|---|------|
| 1 | Table of Contents | |
| 2 | Chapter 5: Food Security | 5-1 |
| 3 | <i>Executive summary</i> | 5-5 |
| 4 | 5.1 <i>Framing and context</i> | 5-8 |
| 5 | 5.1.1 Food security and insecurity, the food system, and climate change | 5-8 |
| 6 | 5.1.1.1 Food security as an outcome of the food system | 5-9 |
| 7 | 5.1.1.2 Effects of climate change on food security | 5-10 |
| 8 | 5.1.2 Status of the food system, food insecurity, and malnourishment..... | 5-12 |
| 9 | 5.1.2.1 Trends in the global food system | 5-12 |
| 10 | 5.1.2.2 Food insecurity status and trends | 5-14 |
| 11 | 5.1.3 Climate change, gender, and equity | 5-16 |
| 12 | Box 5.1 Gender, food security, and climate change | 5-17 |
| 13 | 5.1.4 Food systems in AR5, SR15, and the Paris Agreement | 5-18 |
| 14 | 5.1.4.1 Food systems in AR5 and SR15 | 5-19 |
| 15 | 5.1.4.2 Food systems and the Paris Agreement | 5-20 |
| 16 | 5.1.4.3 Charting the future of food security | 5-21 |
| 17 | 5.2 <i>Impacts of climate change on food systems</i> | 5-22 |
| 18 | 5.2.1 Climate drivers important to food security | 5-22 |
| 19 | 5.2.1.1 Short-lived climate pollutants | 5-23 |
| 20 | 5.2.2 Climate change impacts on food availability | 5-24 |
| 21 | 5.2.2.1 Impacts on crop production..... | 5-24 |
| 22 | 5.2.2.2 Impacts on livestock production systems..... | 5-28 |
| 23 | 5.2.2.3 Impacts on pests and diseases | 5-32 |
| 24 | 5.2.2.4 Impacts on pollinators..... | 5-33 |
| 25 | 5.2.2.5 Impacts on aquaculture | 5-34 |
| 26 | 5.2.2.6 Impacts on smallholder farming systems | 5-35 |
| 27 | 5.2.3 Climate change impacts on access | 5-37 |
| 28 | 5.2.3.1 Impacts on prices and risk of hunger | 5-37 |
| 29 | 5.2.3.2 Impacts on land use..... | 5-39 |
| 30 | 5.2.4 Climate change impacts on food utilisation | 5-39 |
| 31 | 5.2.4.1 Impacts on food safety and human health..... | 5-39 |
| 32 | 5.2.4.2 Impacts on food quality..... | 5-40 |
| 33 | 5.2.5 Climate change impacts on food stability | 5-42 |
| 34 | 5.2.5.1 Impacts of extreme events..... | 5-42 |
| 35 | 5.2.5.2 Food aid | 5-42 |
| 36 | 5.3 <i>Adaptation options, challenges, and opportunities</i> | 5-43 |
| 37 | 5.3.1 Challenges and opportunities | 5-43 |
| 38 | Box 5.2 Sustainable solutions for food systems and climate change in Africa..... | 5-44 |
| 39 | 5.3.2 Adaptation framing and key concepts..... | 5-45 |
| 40 | 5.3.2.1 Autonomous, incremental, and transformational adaptation | 5-45 |
| 41 | 5.3.2.2 Risk management..... | 5-47 |
| 42 | 5.3.2.3 Role of agroecology and diversification | 5-48 |
| 43 | Box 5.3 Climate change and indigenous food systems in the Hindu-Kush Himalayan Region | 5-49 |
| 44 | 5.3.2.4 Role of cultural values | 5-51 |
| 45 | 5.3.3 Supply-side adaptation..... | 5-52 |
| 46 | 5.3.3.1 Crop production | 5-52 |
| 47 | 5.3.3.2 Livestock production systems | 5-53 |

| | | | |
|----|----------------------|---|-------------|
| 1 | 5.3.3.3 | Aquaculture, fisheries, and agriculture interactions..... | 5-53 |
| 2 | 5.3.3.4 | Transport and storage..... | 5-54 |
| 3 | 5.3.3.5 | Trade and processing | 5-54 |
| 4 | 5.3.4 | Demand-side adaptation..... | 5-55 |
| 5 | 5.3.5 | Institutional measures | 5-56 |
| 6 | 5.3.5.1 | Global initiatives..... | 5-56 |
| 7 | 5.3.5.2 | National policies | 5-57 |
| 8 | 5.3.5.3 | Community-based adaptation | 5-58 |
| 9 | 5.3.6 | Tools and finance..... | 5-59 |
| 10 | 5.3.6.1 | Early Warning Systems..... | 5-59 |
| 11 | 5.3.6.2 | Financial resources..... | 5-59 |
| 12 | 5.4 | <i>Impacts of food systems on climate change</i> | <i>5-60</i> |
| 13 | 5.4.1 | Greenhouse gas emissions from food systems..... | 5-60 |
| 14 | 5.4.2 | Greenhouse gas emissions from croplands and soils | 5-61 |
| 15 | 5.4.3 | Greenhouse gas emissions from livestock | 5-62 |
| 16 | 5.4.4 | Greenhouse gas emissions from aquaculture | 5-64 |
| 17 | 5.4.5 | Greenhouse gas emissions from inputs, processing, storage, and transport..... | 5-64 |
| 18 | 5.4.6 | Greenhouse gas emissions associated with different diets..... | 5-65 |
| 19 | 5.5 | <i>Mitigation options, challenges, and opportunities.....</i> | <i>5-66</i> |
| 20 | 5.5.1 | Supply-side mitigation options | 5-67 |
| 21 | Box 5.4 | Towards sustainable intensification in South America region | 5-68 |
| 22 | 5.5.1.1 | Greenhouse gas mitigation in croplands and soils | 5-70 |
| 23 | 5.5.1.2 | Greenhouse gas mitigation in livestock systems..... | 5-71 |
| 24 | 5.5.1.3 | Greenhouse gas mitigation in agroforestry | 5-72 |
| 25 | 5.5.1.4 | Integrated approaches to crop and livestock mitigation..... | 5-73 |
| 26 | 5.5.1.5 | Greenhouse gas mitigation in aquaculture | 5-75 |
| 27 | 5.5.1.6 | Cellular agriculture | 5-75 |
| 28 | 5.5.2 | Demand-side mitigation options | 5-76 |
| 29 | 5.5.2.1 | Demand-side scenarios of different diets..... | 5-76 |
| 30 | 5.5.2.2 | Role of dietary preferences | 5-78 |
| 31 | 5.5.2.3 | Uncertainties in demand-side technical mitigation potential | 5-79 |
| 32 | 5.5.2.4 | Insect-based diets | 5-80 |
| 33 | 5.5.2.5 | Food loss and waste, food security, and land use | 5-80 |
| 34 | 5.5.2.6 | Shortening supply chains | 5-82 |
| 35 | 5.6 | <i>Mitigation, Adaptation, Food Security, and Land Use – Synergies, Trade-Offs, and Co-</i> | |
| 36 | <i>Benefits.....</i> | <i>5-83</i> | |
| 37 | 5.6.1 | Land-based carbon dioxide removal (CDR) and bioenergy..... | 5-84 |
| 38 | 5.6.2 | Mitigation, food prices, and food security | 5-86 |
| 39 | 5.6.3 | Environmental and health effects of adopting healthy and sustainable diets..... | 5-89 |
| 40 | 5.6.3.1 | Can dietary shifts provide significant benefits? | 5-90 |
| 41 | 5.6.4 | Sustainable integrated agricultural systems | 5-91 |
| 42 | 5.6.4.1 | Agroecology..... | 5-92 |
| 43 | 5.6.4.2 | Climate-smart agriculture | 5-93 |
| 44 | 5.6.4.3 | Conservation agriculture | 5-94 |
| 45 | 5.6.4.4 | Sustainable intensification | 5-95 |
| 46 | Cross-Chapter Box 6: | Agricultural intensification: land sparing, land sharing and sustainability 5- | |
| 47 | | 96 | |
| 48 | 5.6.5 | Role of urban agriculture | 5-100 |

| | | | |
|----|---------|---|-------|
| 1 | 5.6.6 | Links to the Sustainable Development Goals | 5-102 |
| 2 | 5.7 | <i>Enabling conditions and knowledge gaps</i> | 5-103 |
| 3 | 5.7.1 | Enabling policy environments | 5-104 |
| 4 | 5.7.1.1 | Agriculture and trade policy | 5-104 |
| 5 | 5.7.1.2 | Scope for expanded policies | 5-105 |
| 6 | 5.7.1.3 | Health-related policies and cost savings | 5-108 |
| 7 | 5.7.1.4 | Multiple policy pathways..... | 5-108 |
| 8 | 5.7.2 | Enablers for changing markets and trade | 5-109 |
| 9 | 5.7.2.1 | Capital markets | 5-109 |
| 10 | 5.7.2.2 | Insurance and re-insurance..... | 5-110 |
| 11 | 5.7.3 | Just transitions to sustainability | 5-110 |
| 12 | 5.7.4 | Mobilising knowledge..... | 5-111 |
| 13 | 5.7.4.1 | Indigenous and local knowledge..... | 5-111 |
| 14 | 5.7.4.2 | Citizen science | 5-111 |
| 15 | 5.7.4.3 | Capacity building and education..... | 5-111 |
| 16 | 5.7.5 | Knowledge gaps and key research areas..... | 5-112 |
| 17 | 5.7.5.1 | Impacts and adaptation..... | 5-112 |
| 18 | 5.7.5.2 | Emissions and mitigation..... | 5-113 |
| 19 | 5.7.5.3 | Synergies and trade-offs..... | 5-114 |
| 20 | 5.8 | <i>Future challenges to food security</i> | 5-114 |
| 21 | 5.8.1 | Food price spikes | 5-115 |
| 22 | Box 5.5 | Market drivers and the consequences of extreme weather in 2010-2011..... | 5-116 |
| 23 | 5.8.2 | Migration and conflict..... | 5-117 |
| 24 | 5.8.2.1 | Migration..... | 5-117 |
| 25 | Box 5.6 | Migration in the Pacific region: Impacts of climate change on food security | 5-118 |
| 26 | 5.8.2.2 | Conflict | 5-120 |
| 27 | | <i>Frequently Asked Questions</i> | 5-120 |
| 28 | | <i>References</i> | 5-121 |
| 29 | | | |
| 30 | | | |

1 **Executive summary**

2 **The current food system (production, transport, processing, packaging, storage, retail,**
3 **consumption, loss and waste) feeds the great majority of world population and supports the**
4 **livelihoods of ca. 200 million people.** Since 1961, food supply per capita has increased more than
5 30%, accompanied by greater use of nitrogen fertilisers (increase of about 800%) and water resources
6 for irrigation (increase of more than 100%). However, an estimated 821 million people are currently
7 undernourished, 151 million children under 5 are stunted, 613 million women and girls aged 15 to 49
8 suffer from iron deficiency, and 2 billion adults are overweight or obese. The food system is under
9 pressure from non-climate stressors (e.g., population and income growth, demand for animal-sourced
10 products), and from climate change. These climate and non-climate stresses are impacting the four
11 pillars of food security (availability, access, utilisation, and stability). {5.1.1, 5.1.2}

12 **Observed climate change is already affecting food security through increasing temperatures,**
13 **changing precipitation patterns, and greater frequency of some extreme events (*high***
14 ***confidence*).** Increasing temperatures are affecting agricultural productivity in higher latitudes, raising
15 yields of some crops (maize, cotton, wheat, sugar beets), while yields of others (maize, wheat, barley)
16 are declining in lower-latitude regions. Warming compounded by drying has caused yield declines in
17 parts of Southern Europe. Based on indigenous and local knowledge, climate change is affecting food
18 security in drylands, particularly those in Africa, and high mountain regions of Asia and South
19 America. {5.2.2}

20 **Food security will be increasingly affected by projected future climate change (*high confidence*).**
21 Across SSPs 1, 2, and 3, global crop and economic models projected a 1-29% cereal price increase in
22 2050 due to climate change (RCP 6.0), which would impact consumers globally through higher food
23 prices; regional effects will vary (*high confidence*). Low-income consumers are particularly at risk,
24 with models projecting increases of 1-183 million additional people at risk of hunger across the SSPs
25 compared to a no climate change scenario (*high confidence*). While increased CO₂ is projected to be
26 beneficial for crop productivity at lower temperature increases, it is projected to lower nutritional
27 quality (*high confidence*) (e.g., wheat grown at 546-586 ppm CO₂ has 5.9–12.7% less protein, 3.7–
28 6.5% less zinc, and 5.2–7.5% less iron). Distributions of pests and diseases will change, affecting
29 production negatively in many regions (*high confidence*). Given increasing extreme events and
30 interconnectedness, risks of food system disruptions are growing (*high confidence*). {5.2.3, 5.2.4}

31 **Vulnerability of pastoral systems to climate change is very high (*high confidence*).** Pastoralism is
32 practiced in more than 75% of countries by between 200 and 500 million people, including nomadic
33 communities, transhumant herders, and agro-pastoralists. Impacts in pastoral systems include lower
34 pasture and animal productivity, damaged reproductive function, and biodiversity loss. Pastoral
35 system vulnerability is exacerbated by non-climate factors (land tenure, sedentarisation, changes in
36 traditional institutions, invasive species, lack of markets, and conflicts). {5.2.2}

37 **Fruit and vegetable production, a key component of healthy diets, is also vulnerable to climate**
38 **change (*medium evidence, high agreement*).** Declines in yields and crop suitability are projected
39 under higher temperatures, especially in tropical and semi-tropical regions. Heat stress reduces fruit
40 set and speeds up development of annual vegetables, resulting in yield losses, impaired product
41 quality, and increasing food loss and waste. Longer growing seasons enable a greater number of
42 plantings to be cultivated and can contribute to greater annual yields. However, some fruits and
43 vegetables need a period of cold accumulation to produce a viable harvest, and warmer winters may
44 constitute a risk. {5.2.2}

45 **Food security and climate change have strong gender and equity dimensions (*high confidence*).**
46 Worldwide, women play a key role in food security, although regional differences exist. Climate
47 change impacts vary among diverse social groups depending on age, ethnicity, gender, wealth, and

1 class. Climate extremes have immediate and long-term impacts on livelihoods of poor and vulnerable
2 communities, contributing to greater risks of food insecurity that can be a stress multiplier for internal
3 and external migration (*medium confidence*). {5.2.6} Empowering women and rights-based
4 approaches to decision-making can create synergies among household food security, adaptation, and
5 mitigation. {5.6.4}

6 **Many practices can be optimised and scaled up to advance adaptation throughout the food**
7 **system (*high confidence*)**. Supply-side options include increased soil organic matter and erosion
8 control, improved cropland, livestock, and grazing land management, and genetic improvements for
9 tolerance to heat and drought. Diversification in the food system (e.g., implementation of integrated
10 production systems, broad-based genetic resources, and heterogeneous diets) is a key strategy to
11 reduce risks (*medium confidence*). Demand-side adaptation, such as adoption of healthy and
12 sustainable diets, in conjunction with reduction in food loss and waste, can contribute to adaptation
13 through reduction in additional land area needed for food production and associated food system
14 vulnerabilities. Indigenous and local knowledge can contribute to enhancing food system resilience
15 (*high confidence*). {5.3, 5.6.3 Cross-Chapter Box 6}.

16 **Ca. 25-30% of total GHG emissions are attributable to the food system. These are from**
17 **agriculture and land use, storage, transport, packaging, processing, retail, and consumption**
18 **(*medium confidence*)**. This estimate includes emissions of 10–12% from crop and livestock activities
19 within the farm gate and 8-10% from land use and land use change including deforestation and
20 peatland degradation (*high confidence*); 5–10% is from supply chain activities (*medium confidence*).
21 This estimate includes GHG emissions from food loss and waste. Within the food system, during the
22 period 2007-2016, the major sources of emissions from the supply side were agricultural production,
23 with crop and livestock activities within the farm gate generating respectively $142 \pm 43 \text{ Tg CH}_4 \text{ yr}^{-1}$
24 (*high confidence*) and $8.3 \pm 2.3 \text{ Tg N}_2\text{O yr}^{-1}$ (*high confidence*), and CO_2 emissions linked to relevant
25 land use change dynamics such as deforestation and peatland degradation, generating $4.8 \pm 2.4 \text{ Gt}$
26 $\text{CO}_2 \text{ yr}^{-1}$. Using 100-year GWP values (no climate feedback) from the IPCC AR5, this implies that
27 total GHG emissions from agriculture were $6.2 \pm 1.9 \text{ Gt CO}_2\text{eq yr}^{-1}$, increasing to $11.0 \pm 3.1 \text{ Gt CO}_2\text{eq}$
28 yr^{-1} including relevant land use. Without intervention, these are likely to increase by about 30%–40%
29 by 2050, due to increasing demand based on population and income growth and dietary change (*high*
30 *confidence*). {5.4}

31 **Supply-side practices can contribute to climate change mitigation by reducing crop and**
32 **livestock emissions, sequestering carbon in soils and biomass, and by decreasing emissions**
33 **intensity within sustainable production systems (*high confidence*)**. Total mitigation potential of
34 crop and livestock activities is estimated as 1.5–4.0 $\text{GtCO}_2\text{-eq yr}^{-1}$ by 2030 at prices ranging from 20-
35 100 USD/ tCO_2eq (*high confidence*). Options with large potential for GHG mitigation in cropping
36 systems include soil carbon sequestration (at decreasing rates over time), reductions in N_2O emissions
37 from fertilisers, reductions in CH_4 emissions from paddy rice, and bridging of yield gaps. Options
38 with large potential for mitigation in livestock systems include better grazing land management, with
39 increased net primary production and soil carbon stocks, improved manure management, and higher-
40 quality feed. Reductions in GHG emissions intensity (emissions per unit product) from livestock can
41 support reductions in absolute emissions, provided appropriate governance to limit total production is
42 implemented at the same time (*medium confidence*). {5.5.1}

43 **Consumption of healthy and sustainable diets presents major opportunities for reducing GHG**
44 **emissions from food systems and improving health outcomes (*high confidence*)**. Examples of
45 healthy and sustainable diets are high in coarse grains, pulses, fruits and vegetables, and nuts and
46 seeds; low in energy-intensive animal-sourced and discretionary foods (such as sugary beverages);
47 and with a carbohydrate threshold. Total mitigation potential of dietary changes is estimated as 1.8-
48 3.4 $\text{GtCO}_2\text{eq yr}^{-1}$ by 2050 at prices ranging from 20-100 USD/ tCO_2 (*medium confidence*). This

1 estimate includes reductions in emissions from livestock and soil carbon sequestration on spared land,
2 but co-benefits with health are not taken into account. Mitigation potential of dietary change may be
3 higher, but achievement of this potential at broad scales depends on consumer choices and dietary
4 preferences that are guided by social, cultural, environmental, and traditional factors, as well as
5 income growth. Meat analogues such as imitation meat (from plant products), cultured meat, and
6 insects may help in the transition to more healthy and sustainable diets, although their carbon
7 footprints and acceptability are uncertain. {5.5.2, 5.6.5}

8 **Reduction of food loss and waste could lower GHG emissions and improve food security**
9 **(medium confidence)**. Combined food loss and waste amount to a third of global food production
10 **(high confidence)**. During 2010-2016, global food loss and waste equalled 8–10% of total GHG
11 emissions from food systems **(medium confidence)**; and cost about USD 1 trillion per year (2012
12 prices) **(low confidence)**. Technical options for reduction of food loss and waste include improved
13 harvesting techniques, on-farm storage, infrastructure, and packaging. Causes of food loss (e.g., lack
14 of refrigeration) and waste (e.g., behaviour) differ substantially in developed and developing
15 countries, as well as across regions **(robust evidence, medium agreement)**. {5.5.2}

16 **Agriculture and the food system are key to global climate change responses. Combining supply-**
17 **side actions such as efficient production, transport, and processing with demand-side**
18 **interventions such as modification of food choices, and reduction of food loss and waste, reduces**
19 **GHG emissions and enhances food system resilience (high confidence)**. Such combined measures
20 can enable the implementation of large-scale land-based adaptation and mitigation strategies without
21 threatening food security from increased competition for land for food production and higher food
22 prices. Without combined food system measures in farm management, supply chains, and demand,
23 adverse effects would include increased number of malnourished people and impacts on smallholder
24 farmers **(medium evidence, high agreement)**. Just transitions are needed to address these effects. {5.5,
25 5.6, 5.7}

26 **For adaptation and mitigation throughout the food system, enabling conditions need to be**
27 **created through policies, markets, institutions, and governance (high confidence)**. For adaptation,
28 resilience to increasing extreme events can be accomplished through risk sharing and transfer
29 mechanisms such as insurance markets and index-based weather insurance **(high confidence)**. Public
30 health policies to improve nutrition – such as school procurement, health insurance incentives, and
31 awareness-raising campaigns – can potentially change demand, reduce health-care costs, and
32 contribute to lower GHG emissions **(limited evidence, high agreement)**. Without inclusion of
33 comprehensive food system responses in broader climate change policies, the mitigation and
34 adaptation potentials assessed in this chapter will not be realised and food security will be jeopardised
35 **(high confidence)**. {5.7}

36

37

38

39

40

41

1 **5.1 Framing and context**

2 The current food system (production, transport, processing, packaging, storage, retail, consumption,
3 loss and waste) feeds the great majority of world population and supports the livelihoods of ca. 200
4 million people. Agriculture as an economic activity generates between 1% and 60% of national GDP
5 in many countries, with a world average of about 4% in 2017 (World Bank 2019). Since 1961, food
6 supply per capita has increased more than 30%, accompanied by greater use of nitrogen fertiliser
7 (increase of about 800%) and water resources for irrigation (increase of more than 100%).

8 The rapid growth in agricultural productivity since the 1960s has underpinned the development of the
9 current global food system that is both a major driver of climate change, and increasingly vulnerable
10 to it (from production, transport, and market activities). Given the current food system, the FAO
11 estimates that there is a need to produce about 50% more food by 2050 in order to feed the increasing
12 world population (FAO 2018a). This would engender significant increases in GHG emissions and
13 other environmental impacts, including loss of biodiversity. FAO (2018a) projects that by 2050
14 cropland area will increase 90-325 Mha, between 6-21% more than the 1,567 Mha cropland area of
15 2010, depending on climate change scenario and development pathway (the lowest increase arises
16 from reduced food loss and waste and adoption of more sustainable diets).

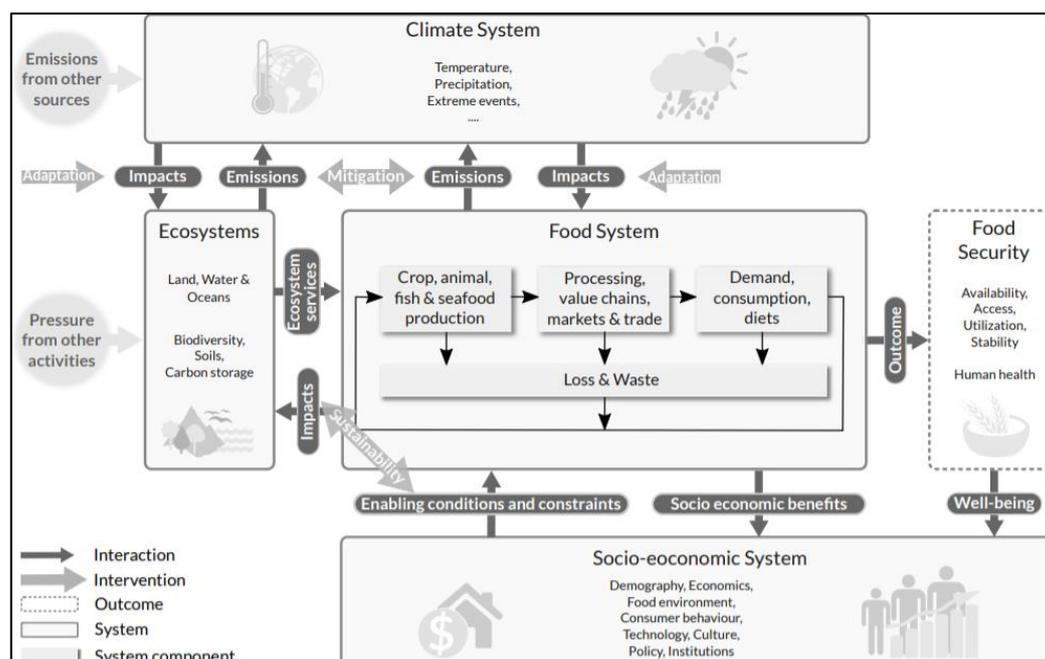
17 Climate change has direct impacts on food systems, food security, and, through the need to mitigate,
18 potentially increases the competition for resources needed for agriculture. Responding to climate
19 change through deployment of land-based technologies for negative emissions based on biomass
20 production would increasingly put pressure on food production and food security through potential
21 competition for land.

22 Using a food system approach, this chapter addresses how climate change affects food security,
23 including nutrition, the options for the food system to adapt and mitigate, synergies and trade-offs
24 among these options, and enabling conditions for their adoption. The chapter assesses the role of
25 incremental and transformational adaptation, and the potential for combinations of supply-side
26 measures such as sustainable intensification (increasing productivity per hectare) and demand-side
27 measures (e.g., dietary change and waste reduction) to contribute to climate change mitigation.

28

29 **5.1.1 Food security and insecurity, the food system, and climate change**

30 The *food system* encompasses all the activities and actors in the production, transport, manufacturing,
31 retailing, consumption, and waste of food, and their impacts on nutrition, health and well-being, and
32 the environment (Figure 5.1).



1
2 **Figure 5.1 Interlinkages between the climate system, food system, ecosystem (land, water and oceans),**
3 **and socio-economic system. These systems operate at multiple scales, both global and regional. Food**
4 **security is an outcome of the food system leading to human well-being, which is also indirectly linked with**
5 **climate and ecosystems through the socio-economic system. Response options for sustainable (S)**
6 **practices, mainly in terms of climate change mitigation (M) and adaptation (A) are represented by grey**
7 **arrows. Adaptation measures can help to reduce negative impacts of climate change on the food system**
8 **and ecosystems. Mitigation measures can reduce greenhouse gas emissions coming from the food system**
9 **and ecosystems.**

10 5.1.1.1 Food security as an outcome of the food system

11 The activities and the actors in the food system leads to outcomes such as food security and generate
12 impacts on the environment. As part of the environmental impacts, food systems are a considerable
13 contributor to greenhouse gas emissions, and thus climate change (Section 5.4). In turn climate
14 change has complex interactions with food systems, leading to food insecurity through impacts on
15 food availability, access, utilisation and stability (Table 5.1; Section 5.2).

16 We take a *food systems lens* in the Special Report on Climate Change and Land (SRCLL) to recognise
17 that demand for and supply of food are interlinked and need to be jointly assessed in order to identify
18 the challenges of mitigation and adaptation to climate change. Outcomes cannot be disaggregated
19 solely to, for example, agricultural production, because the demand for food shapes what is grown,
20 where it is grown, and how much is grown. Thus, greenhouse gas emissions from agriculture result, in
21 large part, from ‘pull’ from the demand side. Mitigation and adaptation involve modifying production,
22 supply chain, and demand practices (through for example dietary choices, market incentives, and
23 trade relationships), so as to evolve a more sustainable and healthy food system.

24 According to FAO (2001a), *food security* is a situation that exists when all people, at all times, have
25 physical, social, and economic access to sufficient, safe, and nutritious food that meets their dietary
26 needs and food preferences for an active and healthy life. “All people at all times” implies the need
27 for equitable and stable food distribution, but it is increasingly recognised that it also covers the need
28 for inter-generational equity, and therefore “sustainability” in food production. “Safe and nutritious
29 food ...for a healthy life” implies that food insecurity can occur if the diet is not nutritious, including
30 when there is consumption of an excess of calories, or if food is not safe, meaning free from harmful
31 substances.

1 A prime impact of food insecurity is *malnourishment* (literally “bad nourishment”) leading to
 2 *malnutrition*, which refers to deficiencies, excesses, or imbalances in a person’s intake of energy
 3 and/or nutrients. As defined by FAO et al. (2018), undernourishment occurs when an individual’s
 4 habitual food consumption is insufficient to provide the amount of dietary energy required to maintain
 5 a normal, active, healthy life. In addition to undernourishment in the sense of insufficient calories
 6 (“hunger”), undernourishment occurs in terms of nutritional deficiencies in vitamins (e.g., Vitamin A)
 7 and minerals (e.g., iron, zinc, iodine), so-called “hidden hunger”. Hidden hunger tends to be present in
 8 countries with high levels of undernourishment (Muthayya et al. 2013), but micronutrient deficiency
 9 can occur in societies with low prevalence of undernourishment. For example, in many parts of the
 10 world teenage girls suffer from iron deficiency (Whitfield et al. 2015) and calcium deficiency is
 11 common in Western-style diets (Aslam and Varani 2016). Food security is related to nutrition, and
 12 conversely food insecurity is related to malnutrition. Not all malnourishment arises from food
 13 insecurity, as households may have access to healthy diets but choose to eat unhealthily, or it may
 14 arise from illness. However, in many parts of the world, poverty is linked to poor diets (FAO et al.
 15 2018). This may be through lack of resources to produce or access food in general, or healthy food, in
 16 particular, as healthier diets are more expensive than diets rich in calories but poor in nutrition (*high*
 17 *confidence*) (see meta-analysis by Darmon and Drewnowski 2015). The relationship between poverty
 18 and poor diets may also be linked to unhealthy “food environments,” with retail outlets in a locality
 19 only providing access to foods of low-nutritional quality (Gamba et al. 2015) – such areas are
 20 sometimes termed “food deserts” (Battersby 2012).

21 Whilst conceptually the definition of food security is clear, it is not straightforward to measure in a
 22 simple way that encompasses all its aspects. Although there are a range of methods to assess food
 23 insecurity, they all have some shortcomings. For example, the UN FAO has developed the Food
 24 Insecurity Experience Scale (FIES), a survey-based tool to measure the severity of overall
 25 households’ inability to access food. While it provides reliable estimates of the prevalence of food
 26 insecurity in a population, it does not reveal whether actual diets are adequate or not with respect to
 27 all aspects of nutrition (see Section 5.1.2.1).

28

29 **5.1.1.2 Effects of climate change on food security**

30 Climate change is projected to negatively impact the four pillars of food security – availability,
 31 access, utilisation and stability – and their interactions (FAO et al. 2018) (*high confidence*). This
 32 chapter assesses recent work since AR5 that has strengthened understanding of how climate change
 33 affects each of these pillars across the full range of food system activities (Table 5.1, Section 5.2).

34 While most studies continue to focus on availability via impacts on food production, more studies are
 35 addressing related issues of access (e.g., impacts on food prices), utilisation (e.g., impacts on
 36 nutritional quality), and stability (e.g., impacts of increasing extreme events) as they are affected by a
 37 changing climate (Bailey et al. 2015). Low-income producers and consumers are likely to be most
 38 affected because of a lack of resources to invest in adaptation and diversification measures (UNCCD
 39 2017; Bailey et al. 2015).

40

41 **Table 5.1 Relationships between food security, the food system, and climate change and guide to chapter.**

| Food security pillar | Examples of observed and projected climate change impacts | Sections | Examples of adaptation and mitigation | Section |
|-----------------------------------|---|---------------------|---------------------------------------|---------|
| Availability <i>Production</i> | Reduced yields in crop and livestock systems | 5.2.2.1, 5.2.2.2 | Development of adaptation practices | 5.3 |

| | | | | |
|--|---|----------------------------------|---|-------------------------|
| <i>of food and its readiness for use through storage, processing, distribution, sale and/or exchange</i> | Reduced yields from lack of pollinators; pests and diseases | 5.2.2.3, 5.2.2.4 | Adoption of new technologies, new and neglected varieties | 5.3.2.3, 5.3.3.1, |
| | Reduced food quality affecting availability (e.g., food spoilage and loss from mycotoxins) | 5.2.4.1, 5.5.2.5 | Enhanced resilience by integrated practices, better food storage | 5.3.2.3, 5.3.3.4, 5.6.4 |
| | Disruptions to food storage and transport networks from change in climate, including extremes | 5.2.5.1, 5.3.3.4, 5.8.1, Box 5.5 | Reduction of demand on by reducing waste, modifying diets | 5.3.4, 5.5.2, 5.7 |
| | | | Closing of crop yield and livestock productivity gaps | 5.6.4.4, 5.7 |
| | | | Risk management, including marketing mechanisms, financial insurance | 5.3.2, 5.7 |
| <i>Access: Ability to obtain food, including effects of price</i> | Yield reductions, changes in farmer livelihoods, limitations on ability to purchase food | 5.2.2.1, 5.2.2.2 | Integrated agricultural practices to build resilient livelihoods | 5.6.4 |
| | Price rise and spike effects on low-income consumers, in particular women and children, due to lack of resources to purchase food | 5.1.3, 5.2.3.1, 5.2.5.1, Box 5.1 | Increased supply chain efficiency (e.g., reducing loss and waste) | 5.3.3, 5.3.4 |
| | Effects of increased extreme events on food supplies, disruption of agricultural trade and transportation infrastructure | 5.8.1 | More climate-resilient food systems, shortened supply chains, dietary change, market change | 5.7 |
| <i>Utilisation Achievement of food potential through nutrition, cooking, health</i> | Impacts on food safety due to increased prevalence of microorganisms and toxins | 5.2.4.1 | Improved storage and cold chains | 5.3.3, 5.3.4 |
| | Decline in nutritional quality resulting from increasing atmospheric CO ₂ | 5.2.4.2 | Adaptive crop and livestock varieties, healthy diets, better sanitation | 5.3.4, 5.5.2, 5.7 |
| | Increased exposure to diarrheal and other infectious diseases due to increased risk of flooding | 5.2.4.1 | | |
| <i>Stability Continuous availability and access to food without disruption</i> | Greater instability of supply due to increased frequency and severity of extreme events; food price rises and spikes; instability of agricultural incomes | 5.2.5, 5.8.1 | Resilience via integrated systems and practices, diversified local agriculture, infrastructure investments, modifying markets and trade, reducing food loss and waste | 5.6.4, 5.7, 5.8.1 |
| | Widespread crop failure contributing to migration and conflict | 5.8.2 | Crop insurance for farmers to cope with extreme events | 5.3.2.2, 5.7 |
| | | | Capacity building to develop resilient systems | 5.3.6, 5.7.4 |

| | | | | |
|---|--|---------------------|--|------------|
| Combined <i>Systemic impacts from interactions of all four pillars</i> | Increasing undernourishment as food system is impacted by climate change | 5.1 | Increased food system productivity and efficiency (e.g., supply side mitigation, reducing waste, dietary change) | 5.5.1, 5.7 |
| | Increasing obesity and ill health through narrow focus on adapting limited number commodity crops | 5.1 | Increased production of healthy food and reduced consumption of energy-intensive products | 5.5.2, 5.7 |
| | Increasing environmental degradation and GHG emissions | Cross-Chapter Box 6 | Development of climate smart food systems by reducing GHG emissions, building resilience, adapting to climate change | 5.3.3, 5.7 |
| | Increasing food insecurity due to competition for land and natural resources (e.g., for land-based mitigation) | 5.6.1 | Governance and institutional responses (including food aid) that take into consideration gender and equity | 5.2.5, 5.7 |

1

2 **5.1.2 Status of the food system, food insecurity, and malnourishment**

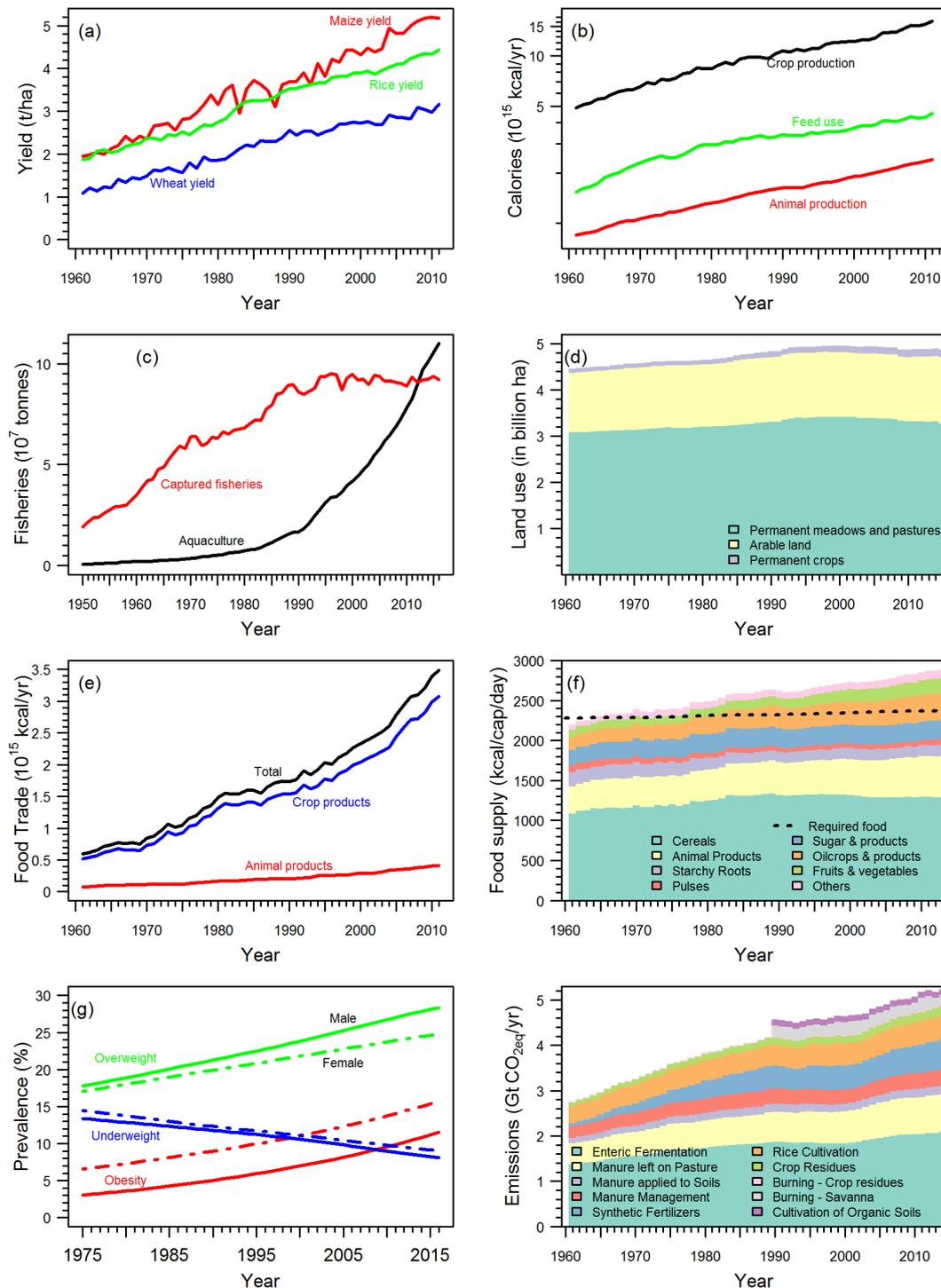
3 **5.1.2.1 Trends in the global food system**

4 Food is predominantly produced on land, with, on average, 83% of the 697 kg of food consumed per
 5 person per year, 93% of the 2884 kcal per day, and 80% of the 81 g of protein eaten per day coming
 6 from terrestrial production in 2013 (FAOSTAT 2018)¹. With increases in crop yields and production
 7 (Figure 5.2), the absolute supply of food has been increasing over the last five decades. Growth in
 8 production of animal-sourced food is driving crop utilisation for livestock feed (FAOSTAT 2018;
 9 Pradhan et al. 2013a). Global trade of crop and animal-sourced food has increased by around 5 times
 10 between 1961 and 2013 (FAOSTAT 2018). During this period, global food availability has increased
 11 from 2200 kcal/cap/day to 2884 kcal/cap/day, making a transition from a food deficit to a food surplus
 12 situation (FAOSTAT 2018; Hiç et al. 2016).

13 The availability of cereals, animal products, oil crops, and fruits and vegetables has mainly grown
 14 (FAOSTAT 2018), reflecting shifts towards more affluent diets. This, in general, has resulted in a
 15 decrease in prevalence of underweight and an increase in prevalence of overweight and obesity
 16 among adults (Abarca-Gómez et al. 2017). During the period 1961-2016, anthropogenic greenhouse
 17 gas emissions associated with agricultural production has grown from 3.1 Gt CO₂-eq yr⁻¹ to 5.8 Gt
 18 CO₂-eq yr⁻¹ (Section 5.4.2, Chapter 2). The increase in emissions is mainly from the livestock sector
 19 (from enteric fermentation and manure left on pasture), use of synthetic fertiliser, and rice cultivation
 20 (FAOSTAT 2018).

21

¹ FOOTNOTE: Does not take into account terrestrial production of feed.



1
2
3
4
5
6
7
8
9
10
11

Figure 5.2 Global trends in (a) yields of maize, rice, and wheat (FAOSTAT 2018) – the top three crops grown in the world; (b) production of crop and animal calories and use of crop calories as livestock feed (FAOSTAT 2018); (c) production from marine and aquaculture fisheries (FishStat 2019); (d) land used for agriculture (FAOSTAT 2018); (e) food trade in calories (FAOSTAT 2018); (f) food supply and required food (i.e., based on human energy requirements for medium physical activities) from 1961–2012 (FAOSTAT 2018; Hiç et al. 2016); (g) prevalence of overweight, obesity and underweight from 1975–2015 (Abarca-Gómez et al. 2017); and (h) GHG emissions for the agriculture sector, excluding land use change (FAOSTAT 2018). For figures (b) and (e), data provided in mass units were converted into calories using nutritive factors (FAO 2001b). Data on emissions due to burning of savanna and cultivation of organic soils is provided only after 1990 (FAOSTAT 2018).

5.1.2.2 Food insecurity status and trends

In addressing food security the dual aspects of malnutrition – under-nutrition and micro-nutrient deficiency, as well as over-consumption, overweight, and obesity – need to be considered (Figure 5.2g and Table 5.2). The UN agencies' *State of Food Security and Nutrition 2018* report (FAO et al. 2018) and the *Global Nutrition Report 2017* (Development Initiatives 2017) summarise the global data. The *State of Food Security* report's estimate for undernourished people on a global basis is 821 million, up from 815 million the previous year and 784 million the year before that. Previous to 2014/2015 the prevalence of hunger had been declining over the last three decades. The proportion of young children (under 5) who are stunted (low height-for-age), has been gradually declining, and was 22% in 2017 compared to 31% in 2012 (150.8 million, down from 165.2 million in 2012). In 2017, 50.5 million children (7.5%) under 5 were wasted (low weight for height). Since 2014, undernutrition has worsened, particularly in parts of sub-Saharan Africa, South-Eastern Asia and Western Asia, and recently Latin America. Deteriorations have been observed most notably in situations of conflict and conflict combined with droughts or floods (FAO et al. 2018).

Regarding micronutrient deficiencies known as 'hidden hunger', reporting suggests a prevalence of one in three people globally (FAO 2013a; von Grebmer et al. 2014; Tulchinsky 2010) (Table 5.2). In the last decades, hidden hunger (measured through proxies targeting iron, vitamin A, and zinc deficiencies) worsened in Africa, while it mainly improved in Asia and Pacific (Ruel-Bergeron et al. 2015). In 2016, 613 million women and girls aged 15 to 49 suffered from iron deficiency (Development Initiatives 2018); in 2013, 28.5% of the global population suffered from iodine deficiency; and in 2005, 33.3% of children under five and 15.3% of pregnant women suffered from vitamin A deficiency, and 17.3% of the global population suffered from zinc deficiency (HLPE 2017).

Table 5.2 Global prevalence of various forms of malnutrition

| | HLPE 2017 (UN) | SOFI 2017 (FAO) | GNR 2017 | SOFI 2018 (FAO) | GNR2018 |
|---------------------------------------|--------------------------------------|-------------------------------|--|--|--|
| Overweight but not obese ^a | 1.3 billion | | 1.93 billion | | 1.34 billion (38,9%) ^c |
| Overweight under five | 41 million | 41 million | 41 million | 38 million | 38 million |
| Obesity ^b | 600 million | 600 million (13%) | 641 million | 672 million | 678 million (13,1%) ^c |
| Undernourishment | 800 million | 815 million | 815 million | 821 million | |
| Stunting under five | 155 million | 155 million | 155 million ^d | 151 million | 151 million ^d (22%) |
| Wasting under five | 52 million | 52 million (8%) | 52 million ^d | 50 million | 51 million ^d (7%) |
| MND (iron) | 19.2% of pregnant women ^e | 33% women of reproductive age | 613 million women and girls aged 15 to 49 ^f | 613 million (32.8%) women and girls aged 15 to 49 ^f | 613 million (32.8%) women and girls aged 15 to 49 ^f |

HLPE: High Level Panel of Experts of the committee of world food security; *SOFI*: The State of Food Security and Nutrition in the World; *GNR*: Global Nutrition Report; *MND*: Micro nutrient deficiency (Iron deficiency for year 2016, uses anemia as a proxy (percentage of pregnant women whose haemoglobin level is less than 110

1 grams per litre at sea level and percentage of non-pregnant women whose haemoglobin level is less than 120
2 grams per litre at sea level).

3 ^aBody mass index between 25-29.9 kg/m²

4 ^bBody mass index greater than 30 kg/m²

5 ^cPrevalence of overweight/obesity among adults (age ≥18) in year 2016. Data from NCD Risk data source.

6 ^dUNICEF WHO Joint Malnutrition;

7 ^eIn 2011

8 ^fAnaemia prevalence in girls and women aged 15 to 49

9

10 Globally, as the availability of inexpensive calories from commodity crops increases, so does per
11 capita consumption of calorie-dense foods (Ng et al. 2014; NCD-RisC 2016a; Abarca-Gómez et al.
12 2017; Doak and Popkin 2017). As a result, in every region of the world, the prevalence of obesity
13 (body mass index >30 kg/m²) and overweight (body mass index range between normality [18.5-24.9]
14 and obesity) is increasing. There are now more obese adults in the world than underweight adults (Ng
15 et al. 2014; NCD-RisC 2016a; Abarca-Gómez et al. 2017; Doak and Popkin 2017). In 2016, around
16 two billion adults were overweight, including 678 million suffering from obesity (NCD-RisC 2016a;
17 Abarca-Gómez et al. 2017). The prevalence of overweight and obesity has been observed in all age
18 groups.

19 Around 41 million children under five years and 340 million children and adolescents aged 5–19
20 years were suffering from overweight or obesity in 2016 (NCD-RisC 2016a; FAO et al. 2017; WHO
21 2015). In many high-income countries, the rising trends in children and adolescents suffering from
22 overweight and obesity have stagnated at high levels; however, these have accelerated in parts of Asia
23 and have very slightly reduced in European and Central Asian lower and middle-income countries
24 (Abarca-Gómez et al. 2017; Doak and Popkin 2017; Christmann et al. 2009).

25 There are associations between obesity and non-communicable diseases such as diabetes, dementia,
26 inflammatory diseases (Saltiel and Olefsky 2017), cardio-vascular disease (Ortega et al. 2016) and
27 some cancers, e.g., of the colon, kidney, and liver (Moley and Colditz 2016). There is a growing
28 recognition of the rapid rise in overweight and obesity on a global basis and its associated health
29 burden created through the non-communicable diseases (NCD-RisC 2016a; HLPE 2017).

30 Analyses reported in FAO et al. (2018) highlight the link between food insecurity, as measured by the
31 FIES scale, and malnourishment (*medium agreement, robust evidence*). This varies by
32 malnourishment measure as well as country (FAO et al. 2018). For example, there is *limited evidence*
33 (*low agreement* but multiple studies) that food insecurity and childhood wasting (i.e., or low weight
34 for height) are closely related, but it is very likely (*high agreement, robust evidence*) that childhood
35 stunting and food insecurity are related (FAO et al. 2018). With respect to adult obesity there is *robust*
36 *evidence, with medium agreement*, that food insecurity, arising from poverty reducing access to
37 nutritious diets, is related to the prevalence of obesity, especially in high-income countries and adult
38 females. An additional meta-analysis (for studies in Europe and North America) also finds a negative
39 relationship between income and obesity, with some support for an effect of obesity causing low
40 income (as well as vice versa) (Kim and von dem Knesebeck 2018).

41 As discussed in Section 5.1.1.1, different methods of assessing food insecurity can provide differential
42 pictures. Of particular note is the spatial distribution of food insecurity, especially in higher-income
43 countries. FAO et al. (2018) reports FIES estimates of severe food insecurity in Africa, Asia and Latin
44 America of 29.8%, 6.9% and 9.8% of the population, respectively, but of 1.4% of the population (i.e.,
45 about 20 million in total; pro rata <5 million for US, <1 million for UK) in Europe and North
46 America. However, in the United States, USDA estimates 40 million people were exposed to varying
47 degrees of food insecurity, from mild to severe (overall prevalence about 12%) (Coleman-Jensen et al.
48 2018). In the UK, estimates from 2017 and 2018 indicate about 4 million adults are moderately to

1 severely food insecure (prevalence 8%) (End Hunger UK 2018; Bates et al. 2017). The UK food bank
2 charity, the Trussell Trust, over a year in 2017/18, distributed 1,332,952 three-day emergency food
3 parcels to people referred to the charity as being in food crisis. Furthermore, a 2003 study in the UK
4 (Schenker 2003) estimated that 40% of adults, and 15% of children, admitted to hospitals were
5 malnourished, and that 70% of undernourishment in the UK was unreported.

6 In total, more than half the world's population are underweight or overweight (NCD-RisC 2017a), so
7 their diets do not provide the conditions for 'an active and healthy life'. This will be more
8 compromised under the impacts of climate change by changing the availability, access, utilisation,
9 and stability of diets of sufficient nutritional quality as shown in Table 5.2 and discussed in detail
10 below (see Section 5.2).

12 **5.1.3 Climate change, gender, and equity**

13 Throughout, the chapter considers many dimensions of gender and equity in regard to climate change
14 and the food system (Box 5.1). Climate change impacts differ among diverse social groups depending
15 on factors such as age, ethnicity, ability/disability, sexual orientation, gender, wealth, and class (*high*
16 *confidence*) (Vincent and Cull 2014; Kaijser and Kronsell 2014). Poverty, along with socio-economic
17 and political marginalisation, cumulatively put women, children and the elderly in a disadvantaged
18 position in coping with the adverse impacts of the changing climate (UNDP 2013; Skoufias et al.
19 2011). The contextual vulnerability of women is higher due to their differentiated relative power,
20 roles, and responsibilities at the household and community levels (Bryan and Behrman 2013; Nelson
21 et al. 2002). They often have a higher reliance on subsistence agriculture, which will be severely
22 impacted by climate change (Aipira et al. 2017).

23 Through impacts on food prices (section 5.2.3.1) poor people's food security is particularly
24 threatened. Decreased yields can impact nutrient intake of the poor by decreasing supplies of highly
25 nutritious crops and by promoting adaptive behaviours that may substitute crops that are resilient but
26 less nutritious (Thompson et al. 2012; Lobell and Burke 2010). In Guatemala, food prices and poverty
27 have been correlated with lower micronutrient intakes (Iannotti et al. 2012). In the developed world,
28 poverty is more typically associated with calorically-dense but nutrient-poor diets, obesity,
29 overweight, and other related diseases (Darmon and Drewnowski 2015).

30 Rural areas are especially affected by climate change (Dasgupta et al. 2014), through impacts on
31 agriculture-related livelihoods and rural income (Mendelsohn et al. 2007) and through impacts on
32 employment. Jessoe et al. (2018) using a 28-year panel on individual employment in rural Mexico,
33 found that years with a high occurrence of heat lead to a reduction in local employment by up to 1.4%
34 with a medium emissions scenario, particularly for wage work and non-farm labour, with impacts on
35 food access. Without employment opportunities in areas where extreme poverty is prevalent, people
36 may be forced to migrate, exacerbating potential for ensuing conflicts (FAO 2018a).

37 Finally, climate change can affect human health in other ways that interact with food utilisation. In
38 many parts of the world where agriculture relies still on manual labour, projections are that heat stress
39 will reduce the hours people can work, and increase their risk (Dunne et al. 2013). For example,
40 Takakura et al (2017) estimates that under RCP8.5, the global economic loss from people working
41 shorter hours to mitigate heat loss may be 2.4–4% of GDP. Furthermore, as discussed by (Watts et al.
42 2018); people's nutritional status interacts with other stressors and affects their susceptibility to ill
43 health (the "utilisation pillar" of food security): so food-insecure people are more likely to be
44 adversely affected by extreme heat, for example.

45 In the case of food price hikes, those more vulnerable are more affected (Uraguchi 2010), especially
46 in urban areas (Ruel et al. 2010), where livelihood impacts are particularly severe for the individuals

1 and groups that have scarce resources or are socially isolated (Revi et al. 2014; Gasper et al. 2011)
2 (*high confidence*). These people often lack power and access to resources, adequate urban services
3 and functioning infrastructure. As climate events become more frequent and intense, this can increase
4 the scale and depth of urban poverty (Rosenzweig et al. 2018b). Urban floods and droughts may result
5 in water contamination increasing the incidence of diarrhoeal illness in poor children (Bartlett 2008).
6 In the near destruction of New Orleans by Hurricane Katrina, about 40,000 jobs were lost (Rosemberg
7 2010).

9 **Box 5.1 Gender, food security, and climate change**

10 Differentiated impacts, vulnerability, risk perception, behaviours and coping strategies for climate
11 change related to food security derive from cultural (gendered) norms, that is, the behaviours, tasks,
12 and responsibilities a society defines as “male” or “female”, and the differential gendered access to
13 resources (Paris and Rola-Rubzen 2018; Aberman and Tirado 2014; Lebel et al. 2014; Bee 2016). In
14 many rural areas women often grow most of the crops for domestic consumption and are primarily
15 responsible for storing, processing, and preparing food; handling livestock; gathering food, fodder and
16 fuelwood; managing domestic water supply; and providing most of the labour for post-harvest
17 activities (FAO 2011a). They are mostly impacted through increased hardship, implications for
18 household roles, and subsequent organisational responsibilities (Boetto and McKinnon 2013; Jost et
19 al. 2016). Water scarcity can particularly affect women because they need to spend more time and
20 energy to collect water, where they may be more exposed to physical and sexual violence (Sommer et
21 al. 2015; Aipira et al. 2017). They may be forced to use unsafe water in the household increasing risk
22 of water-borne diseases (Parikh 2009). Climate change also has differentiated gendered impacts on
23 livestock-holders food security (McKune et al. 2015; Ongoro and Ogara 2012; Fratkin et al. 2004)
24 (See Supplementary Material Table SM5.1).

25 **Gender dimensions of the four pillars**

26 Worldwide, women play a key role in food security (World Bank 2015) and the four pillars of food
27 security have strong gender dimensions (Thompson 2018). In terms of food availability, women tend
28 to have less access to productive resources, including land, and thus less capacity to produce food
29 (Cross-chapter box 11: Gender in Chapter 7).

30 In terms of food access, gendered norms in how food is divided at mealtimes may lead to smaller food
31 portions for women and girls. Women’s intra-household inequity limits their ability to purchase food;
32 limitations also include lack of women’s mobility impacting trips to the market and lack of decision-
33 making within the household (Ongoro and Ogara 2012; Mason et al. 2017; Riley and Dodson 2014).

34 In terms of food utilisation, men, women, children and the elderly have different nutritional needs
35 (e.g., during pregnancy or breast-feeding).

36 In terms of stability, women are more likely to be disproportionately affected by price spikes
37 (Vellakkal et al. 2015; Arndt et al. 2016; Hossain and Green 2011; Darnton-Hill and Cogill 2010;
38 Cohen and Garrett 2010; Kumar and Quisumbing 2013) because when food is scarce women reduce
39 food consumption relative to other family members, although these norms vary according to age,
40 ethnicity, culture, region, and social position, as well as by location in rural or urban areas (Arora-
41 Jonsson 2011; Goh 2012; Niehof 2016; Ongoro and Ogara 2012).

42 **Integrating gender into adaptation**

43 Women have their own capabilities to adapt to climate change. In the Pacific Islands, women hold
44 critical knowledge on where or how to find clean water; which crops to grow in a flood or a drought
45 season; how to preserve and store food and seeds ahead of approaching storms, floods or droughts;

1 and how to carry their families through the recovery months. They also play a pivotal role in
2 managing household finances and investing their savings in education, health, livelihoods, and other
3 activities that assist their families to adapt and respond to climate effects (Aipira et al. 2017).
4 Decreasing women’s capacity to adapt to the impacts of climate change also decreases that of the
5 household (Bryan and Behrman 2013).

6 However, gender norms and power inequalities also shape the ability of men, women, boys, girls and
7 the elderly to adapt to climate risks (Rossi and Lambrou 2008). For example, women pastoralists in
8 the Samburu district of Kenya cannot make decisions affecting their lives, limiting their adaptive
9 capacity (Ongoro and Ogara 2012).

10 Participation in decision-making and politics, division of labour, resource access and control, and
11 knowledge and skills (Nelson and Stathers 2009) are some of the barriers to adaptation. Women’s
12 adaptive capacity is also diminished because their work often goes unrecognised (Rao 2005; Nelson
13 and Stathers 2009). Many of women’s activities are not defined as “economically active employment”
14 in national accounts (FAO 2011a). This non-economic status of women’s activities implies that they
15 are not included in wider discussions of priorities or interventions for climate change. Their
16 perspectives and needs are not met; and thus, interventions, information, technologies, and tools
17 promoted are potentially not relevant, and even can increase discrimination (Alston 2009; Edvardsson
18 Björnberg and Hansson 2013; Huynh and Resurreccion 2014).

19 Where gender-sensitive policies to climate change may exist, effective implementation in practice of
20 gender equality and empowerment may not be achieved on the ground due to lack of technical
21 capacity, financial resources and evaluation criteria, as shown in the Pacific Islands (Aipira et al.
22 2017). Thus, corresponding institutional frameworks that are well-resourced, coordinated, and
23 informed are required, along with adequate technical capacity within government agencies, NGOs and
24 project teams, to strength collaboration and promote knowledge sharing (Aipira et al. 2017).

25 **Women’s empowerment: Synergies among adaptation, mitigation, and food security**

26 Empowered and valued women in their societies increases their capacity to improve food security
27 under climate change, make substantial contributions to their own well-being, to that of their families
28 and of their communities (Langer et al. 2015; Ajani et al. 2013; Alston 2014) (*high confidence*).
29 Women’s empowerment includes economic, social and institutional arrangements and may include
30 targeting men in integrated agriculture programs to change gender norms and improve nutrition (Kerr
31 et al. 2016). Empowerment through collective action and groups-based approaches in the near-term
32 has the potential to equalise relationships on the local, national and global scale (Ringler et al. 2014).
33 Empowered women are crucial to creating effective synergies among adaptation, mitigation, and food
34 security.

35 In Western Kenya, widows in their new role as main livelihood providers invested in sustainable
36 innovations like rainwater harvesting systems and agroforestry (this can serve as both adaptation and
37 mitigation), and worked together in formalised groups of collective action (Gabrielsson and Ramasar
38 2013) to ensure food and water security. In Nepal, women’s empowerment had beneficial outcomes in
39 maternal and children nutrition, reducing the negative effect of low production diversity (Malapit et
40 al. 2015). Integrated nutrition and agricultural programs have increased women’s decision-making
41 power and control over home gardens in Burkina Faso (van den Bold et al. 2015) with positive
42 impacts on food security.

44 **5.1.4 Food systems in AR5, SR15, and the Paris Agreement**

45 Food, and its relationship to the environment and climate change, has grown in prominence since the
46 Rio Declaration in 1992, where food production is Chapter 14 of Agenda 21, to the Paris Agreement

1 of 2015, which includes the need to ensure food security under the threat of climate change on its first
2 page. This growing prominence of food is reflected in recent IPCC reports, including its Fifth
3 Assessment Report (IPCC 2014a) and the Special Report on Global Warming of 1.5°C (SR15) (IPCC
4 2018a).

5

6 **5.1.4.1 Food systems in AR5 and SR15**

7 The IPCC Working Group (WG) II AR5 chapter on Food Security and Food Production Systems
8 broke new ground by expanding its focus beyond the effects of climate change primarily on
9 agricultural production (crops, livestock and aquaculture) to include a food systems approach as well
10 as directing attention to undernourished people (Porter et al. 2014). However, it focused primarily on
11 food production systems due to the prevalence of studies on that topic (Porter et al. 2017). It
12 highlighted that a range of potential adaptation options exist across all food system activities, not just
13 in food production, and that benefits from potential innovations in food processing, packaging,
14 transport, storage, and trade were insufficiently researched at that time.

15 The IPCC WG III AR5 chapter on Agriculture, Forestry and Other Land Use (AFOLU) assessed
16 mitigation potential considering not only the supply, but also the demand side of land uses, by
17 consideration of changes in diets; it also included food loss and waste (Smith et al. 2014). AR5
18 focused on crop and livestock activities within the farm gate and land use and land use change
19 dynamics associated with agriculture. It did not take a full food system approach to emissions
20 estimates that includes processing, transport, storage, and retail.

21 The IPCC WG II AR5 Rural Areas chapter (Revi et al. 2014) found that farm households in
22 developing countries are vulnerable to climate change due to socio-economic characteristics and non-
23 climate stressors, as well as climate risks (Dasgupta et al. 2014). They also found that a wide range of
24 on-farm and off-farm climate change adaptation measures are already being implemented and that the
25 local social and cultural context played a prominent role in the success or failure of different
26 adaptation strategies for food security, such as trade, irrigation or diversification. The IPCC WG II
27 AR5 Urban Areas chapter found that food security of people living in cities was severely affected by
28 climate change through reduced supplies, including urban-produced food, and impacts on
29 infrastructure, as well as a lack of access to food. Poor urban dwellers are more vulnerable to rapid
30 changes of food prices due to climate change.

1 Many climate change response options in IPCC WG II and WG III AR5 (IPCC 2014b) address
2 incremental adaptation or mitigation responses separately rather than being inclusive of more
3 systemic or transformational changes in multiple food systems that are large-scale, in depth, and
4 rapid, requiring social, technological, organisational and system responses (Rosenzweig and Solecki
5 2018; Mapfumo et al. 2017; Termeer et al. 2017). In many cases, transformational change will require
6 integration of resilience and mitigation across all parts of the food system including production,
7 supply chains, social aspects, and dietary choices. Further, these transformational changes in the food
8 system need to encompass linkages to ameliorative responses to land degradation (see Chapter 4),
9 desertification (see Chapter 3), and declines in quality and quantity of water resources throughout the
10 food-energy-water nexus (Chapter 2; Section 5.7).

11 The IPCC Special Report on Global Warming of 1.5°C found that climate-related risks to food
12 security are projected to increase with global warming of 1.5°C and increase further with 2°C (IPCC
13 2018a).

14

15 **5.1.4.2 Food systems and the Paris Agreement**

16 To reach the temperature goal put forward in the Paris Agreement of limiting warming to well below
17 2°C, and pursuing efforts to limit warming to 1.5°C, representatives from 196 countries signed the
18 United Nations Framework Convention on Climate Change (UNFCCC) Paris Agreement (UNFCCC
19 2015) in December 2015. The Agreement put forward a temperature target of limiting warming to
20 well below 2°C, and pursuing efforts to limit warming to 1.5°C. Under the Paris Agreement, Parties
21 are expected to put forward their best efforts through nationally determined contributions (NDCs) and
22 to strengthen these efforts in the years ahead. Article 2 of the Agreement makes clear the agreement is
23 within “the context of sustainable development” and states actions should be “in a manner that does
24 not threaten food production” to ensure food security.

25 Many countries have included food systems in their mitigation and adaptation plans as found in their
26 NDCs for the Paris Agreement (Rosenzweig et al. 2018a). Richards et al. (2015) analysed 160 Party
27 submissions and found that 103 include agricultural mitigation; of the 113 Parties that include
28 adaptation in their NDCs, almost all (102) include agriculture among their adaptation priorities. There
29 is much attention to conventional agricultural practices that can be climate-smart and sustainable (e.g.,
30 crop and livestock management), but less to the enabling services that can facilitate uptake (e.g.,
31 climate information services, insurance, credit). Considerable finance is needed for agricultural
32 adaptation and mitigation by least developed countries – in the order of USD 3 billion annually for
33 adaptation and USD 2 billion annually for mitigation, which may be an underestimate due to a small
34 sample size (Richards et al. 2015). On the mitigation side, none of the largest agricultural emitters
35 included sector-specific contributions from the agriculture sector in their NDCs, but most included
36 agriculture in their economy-wide targets (Richards et al. 2018).

37 *Carbon dioxide removal (CDR)*. A key aspect regarding the implementation of measures to achieve
38 the Paris Agreement goals involves measures related to carbon dioxide removal (CDR) through
39 bioenergy (Sections 5.5 and 5.6). To reach the temperature target put forward of limiting warming to
40 well below 2°C, and pursuing efforts to limit warming to 1.5°C, large investments and abrupt changes
41 in land use will be required to advance bioenergy with carbon capture and sequestration (BECCS),
42 afforestation and reforestation (AR), and biochar technologies. Existing scenarios estimate the global
43 area required for BECCS alone to help limit warming to 1.5°C in the range of 109-990 Mha, most
44 commonly around 380-700 Mha.

45 Most scenarios assume very rapid deployment between 2030 and 2050, reaching rates of expansion in
46 land use in 1.5°C scenarios exceeding 20 M ha yr⁻¹, which are unprecedented for crops and forestry
47 reported in the FAO database from 1961. Achieving the 1.5 °C target would thus result in major

1 competing demands for land between climate change mitigation and food production, with cascading
2 impacts on food security.

3 This chapter assesses how the potential conflict for land could be alleviated by sustainable
4 intensification to produce food with a lower land footprint (Section 5.6, Cross-Chapter Box 6:
5 Agricultural intensification). To accomplish this, farmers would need to produce the same amount of
6 food with lower land requirement, which depends on technology, skills, finance, and markets.
7 Achieving this would also rely on demand-side changes including dietary choices that enable
8 reduction of the land footprint for food production while still meeting dietary needs. Transitions
9 required for such transformative changes in food systems are addressed in Section 5.7.

10

11 **5.1.4.3 Charting the future of food security**

12 This chapter utilises the common framework of the Representative Concentration Pathways (RCPs)
13 and the Shared Socio-economic Pathways (SSPs) (Popp et al. 2017; Riahi et al. 2017; Doelman et al.
14 2018) to assess the impacts of future GHG emissions, mitigation measures, and adaptation on food
15 security (See Cross-Chapter Box 1: Scenarios in Chapter 1, Section 5.2 and 5.6).

16 New work utilising these scenario approaches has shown that the food system externalises costs onto
17 human health and the environment (Springmann et al. 2018a; Swinburn et al. 2019; Willett et al.
18 2019), leading to calls for transforming the food system to deliver better human and sustainability
19 outcomes (Willett et al. 2019; IAP 2018; Development Initiatives 2018; Lozano et al. 2018). Such a
20 transformation could be an important lever to address the complex interactions between climate
21 change and food security. Through acting on mitigation and adaptation in regard to both food demand
22 and food supply we assess the potential for improvements to both human health and the Sustainable
23 Development Goals (Section 5.6).

24 This chapter builds on the food systems and scenario approaches followed by AR5 and its focus on
25 climate change and food security, but new work since AR5 has extended beyond production to how
26 climate change interacts with the whole food system. The analysis of climate change and food
27 insecurity has expanded beyond undernutrition to include the overconsumption of unhealthy mass-
28 produced food high in sugar and fat, which also threatens health in different but highly damaging
29 ways and the role of dietary choices and consumption in greenhouse gas emissions. It focused on
30 land-based food systems, though highlighting in places the contributions of freshwater and marine
31 production.

32 The chapter assesses new work on the observed and projected effects of CO₂ concentrations on the
33 nutritional quality of crops (Section 5.2.4.2) and emphasises the role of extreme climate events
34 (Section 5.2.5.1), social aspects including gender and equity (Box 5.1. and Cross-chapter Box 11:
35 Gender in Chapter 7), and dietary choices (Section 5.4.6, 5.5.2). Other topics with considerable new
36 literature include impacts on smallholder farming systems (Section 5.2.2.6), food loss and waste
37 (Section **Error! Reference source not found.**), and urban and peri-urban agriculture (Section 5.6.5).
38 he chapter explores the potential competing demands for land that mitigation measures to achieve
39 temperature targets may engender, with cascading impacts on food production, food security, and
40 farming systems (Section 5.6), and the enabling conditions for achieving the mitigation and adaptation
41 in equitable and sustainable ways (Section 5.7). Section 5.8 presents challenges to future food
42 security, including food price spikes, migration, and conflict.

43

5.2 Impacts of climate change on food systems

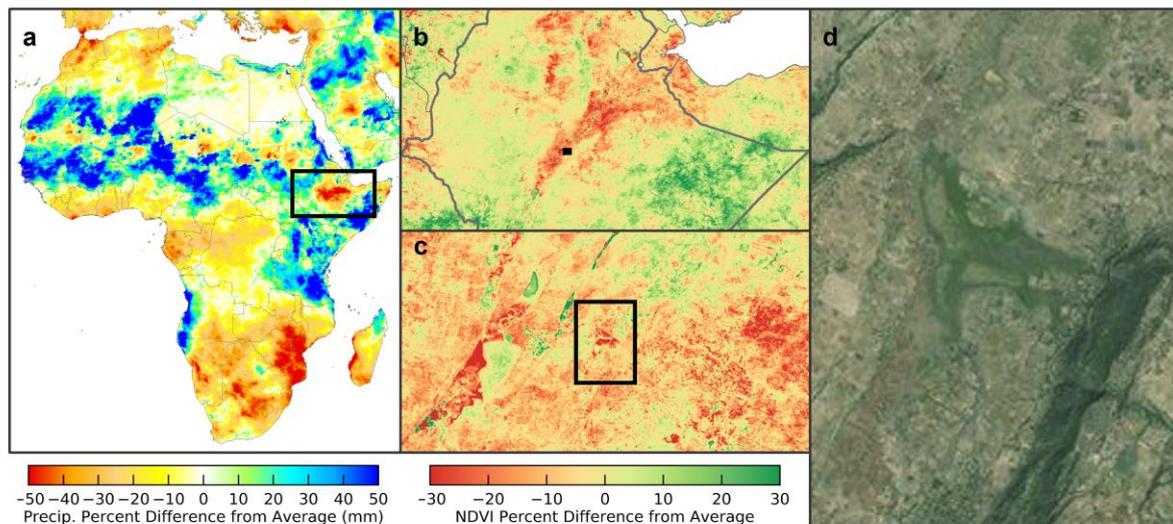
There are many routes by which climate change can impact food security and thus human health (Watts et al. 2018; Fanzo et al. 2017). One major route is via climate change affecting the amount of food, both from direct impacts on yields (Section 5.2.2.1) and indirect effects through climate change's impacts on water availability and quality, pests and diseases (Section 5.2.2.3), and pollination services (Section 5.2.2.4). Another route is via changing CO₂ in the atmosphere, affecting biomass and nutritional quality (Section 5.2.4.2). Food safety risks during transport and storage can also be exacerbated by changing climate (Section 5.2.4.1).

Further, the direct impacts of changing weather can affect human health through the agricultural workforce's exposure to extreme temperatures (Section 5.2.5.1). Through changing metabolic demands and physiological stress for people exposed to extreme temperatures, there is also the potential for interactions with food availability: people may require more food to cope, whilst at the same time being impaired from producing it (Watts et al. 2018). All these factors have the potential to alter both physical health as well as cultural health, through changing the amount, safety and quality of food available for individuals within their cultural context.

This section assesses recent literature on climate change impacts on the four pillars of food security: availability (Section 5.2.2), access (Section 5.2.3), utilisation (Section 5.2.4), and stability (Section 5.2.5). It considers impacts on the food system from climate changes that are already taking place and how impacts are projected to occur in the future. See Supplementary Material Section SM5.2 for discussion of detection and attribution and improvement in projection methods.

5.2.1 Climate drivers important to food security

Climate drivers relevant to food security and food systems include temperature-related, precipitation-related, and integrated metrics that combine these and other variables. These are projected to affect many aspects of the food security pillars (FAO 2018b) (see Supplementary Material Table SM5.2 and Chapter 6 for assessment of observed and projected climate impacts). Climate drivers relevant to food production and availability may be categorised as modal climate changes (e.g., shifts in climate envelopes causing shifts in cropping varieties planted), seasonal changes (e.g., warming trends extending growing seasons), extreme events (e.g., high temperatures affecting critical growth periods, flooding/droughts), and atmospheric conditions (e.g., CO₂ concentrations, short-lived climate pollutants (SLCPs), and dust). Water resources for food production will be affected through changing rates of precipitation and evaporation, ground water levels, and dissolved oxygen content (Cruz-Blanco et al. 2015; Sepulcre-Canto et al. 2014; Huntington et al. 2017; Schmidtko et al. 2017). Potential changes in major modes of climate variability can also have widespread impacts such as occurred during late 2015 to early 2016 when a strong El Niño contributed to regional shifts in precipitation in the Sahel region. Significant drought across Ethiopia resulted in widespread crop failure and more than 10 million people in Ethiopia required food aid (U.S. Department of State 2016; Huntington et al. 2017) (see Figure 5.3).



1
2 **Figure 5.3** Precipitation anomaly and vegetation response in Eastern Africa. (a) Sep 2015–Feb 2016
3 **Climate Hazards Group Infrared Precipitation with Station (CHIRPS) precipitation anomaly over Africa**
4 **relative to the 1981–2010 average shows that large areas of Ethiopia received less than half of normal**
5 **precipitation. Consequently, widespread impacts to agricultural productivity, especially within pastoral**
6 **regions, were present across Ethiopia as evidenced by (d) reduced greenness in remote sensing images. (b)**
7 **MODIS NDVI anomalies for Sep 2015–Feb 2016 relative to 2000–2015 average are shown for the inset**
8 **box in (a). (c) Landsat NDVI anomalies for Sep 2015–Feb 2016 relative to 2000–2015 average are shown**
9 **for the inset box in (b) (Huntington et al. 2017).**

10 Other variables that affect agricultural production, processing, and/or transport are solar radiation,
11 wind, humidity, and (in coastal areas) salinisation and storm surge (Mutahara et al. 2016; Myers et al.
12 2017). Extreme climate events resulting in inland and coastal flooding, can affect the ability of people
13 to obtain and prepare food (Rao et al. 2016; FAO et al. 2018). For direct effects of atmospheric CO₂
14 concentrations on crop nutrient status see Section 5.2.4.2.

15 16 **5.2.1.1 Short-lived climate pollutants**

17 The important role of short-lived climate pollutants such as ozone and black carbon is increasingly
18 emphasised since they affect agricultural production through direct effects on crops and indirect
19 effects on climate (Emberson et al. 2018; Lal et al. 2017; Burney and Ramanathan 2014; Ghude et al.
20 2014) (see Chapters 2 and 4). Ozone causes damage to plants through damages to cellular metabolism
21 that influence leaf-level physiology to whole-canopy and root-system processes and feedbacks; these
22 impacts affect leaf-level photosynthesis senescence and carbon assimilation, as well as whole-canopy
23 water and nutrient acquisition and ultimately crop growth and yield (Emberson et al. 2018). Using
24 atmospheric chemistry and a global integrated assessment model, Chuwah et al. (2015) found that
25 without a large decrease in air pollutant emissions, high ozone concentration could lead to an increase
26 in crop damage of up to 20% in agricultural regions in 2050 compared to projections in which
27 changes in ozone are not accounted for. Higher temperatures are associated with higher ozone
28 concentrations; C3 crops are sensitive to ozone (e.g., soybeans, wheat, rice, oats, green beans,
29 peppers, and some types of cottons) and C4 crops are moderately sensitive (Backlund et al. 2008).

30 Methane increases surface ozone which augments warming-induced losses and some quantitative
31 analyses now include climate, long-lived (CO₂) and multiple short-lived pollutants (CH₄, O₃)
32 simultaneously (Shindell et al. 2017; Shindell 2016). Reduction of tropospheric ozone and black
33 carbon can avoid premature deaths from outdoor air pollution and increases annual crop yields
34 (Shindell et al. 2012). These actions plus methane reduction can influence climate on shorter time

1 scales than those of carbon dioxide–reduction measures. Implementing them substantially reduces the
2 risks of crossing the 2°C threshold and contributes to achievement of the SDGs (Haines et al. 2017;
3 Shindell et al. 2017).

5 5.2.2 Climate change impacts on food availability

6 Climate change impacts food availability through its effect on the production of food and its storage,
7 processing, distribution, and exchange.

9 5.2.2.1 Impacts on crop production

10 **Observed impacts.** Since AR5, there have been further studies that document impacts of climate
11 change on crop production and related variables (See Supplementary Material Table SM5.3). There
12 have been also a few studies that demonstrate a strengthening relationship between observed climate
13 variables and crop yields that indicate future expected warming will have severe impacts on crop
14 production (Mavromatis 2015; Innes et al. 2015). At the global scale, Iizumi et al. (2018) used a
15 counterfactual analysis and found that climate change between 1981-2010 has decreased global mean
16 yields of maize, wheat, and soybeans by 4.1, 1.8 and 4.5%, respectively, relative to preindustrial
17 climate, even when CO₂ fertilisation and agronomic adjustments are considered. Uncertainties (90%
18 probability interval) in the yield impacts are -8.5 to +.5% for maize, -7.5 to +4.3% for wheat, and -8.4
19 to -0.5% for soybeans. For rice, no significant impacts were detected. This study suggests that climate
20 change has modulated recent yields on the global scale and led to production losses, and that
21 adaptations to date have not been sufficient to offset the negative impacts of climate change,
22 particularly at lower latitudes.

23 Dryland settlements are perceived as vulnerable to climate change with regard to food security,
24 particularly in developing countries; such areas are known to have low capacities to cope effectively
25 with decreasing crop yields (Shah et al. 2008; Nellemann et al. 2009). This is of concern because
26 drylands constitute over 40% of the earth’s land area, and are home to 2.5 billion people (FAO et al.
27 2011).

28 *Australia.* In Australia, declines in rainfall and rising daily maximum temperatures based on
29 simulations of 50 sites caused water-limited yield potential to decline by 27% from 1990 to 2015,
30 even though elevated atmospheric CO₂ concentrations had a positive effect (Hochman et al. 2017). In
31 New South Wales, high-temperature episodes during the reproduction stage of crop growth were
32 found to have negative effects on wheat yields, with combinations of low rainfall and high
33 temperatures being the most detrimental (Innes et al. 2015).

34 *Asia.* There are numerous studies demonstrating that climate change is affecting agriculture and food
35 security in Asia. Several studies with remote sensing and statistical data have examined rice areas in
36 northeastern China, the northernmost region of rice cultivation, and found expansion over various
37 time periods beginning in the 1980s, with most of the increase occurring after 2000 (Liu et al. 2014;
38 Wang et al. 2014; Zhang et al. 2017). Rice yield increases have also been found over a similar period
39 (Wang et al. 2014). Multiple factors, such as structural adjustment, scientific and technological
40 progress, and government policies, along with regional warming (1.43°C in the past century)
41 (Fenghua et al. 2006) have been put forward as contributing to the observed expanded rice areas and
42 yield in the region. Shi et al. (2013) indicate that there is a partial match between climate change
43 patterns and shifts in extent and location of the rice-cropping area (2000-2010).

44 There have also been documented changes in winter wheat phenology in Northwest China (He 2015).
45 Consistent with this finding, dates of sowing and emergence of spring and winter wheat were delayed,
46 dates of anthesis and maturity was advanced, and length of reproductive growth period was prolonged

1 from 1981-2011 in a study looking at these crops across China (Liu et al. 2018b). Another study
2 looking in Norwest China demonstrated that there have been changes in the phenology and
3 productivity of spring cotton (Huang and Ji 2015). A study looking at wheat growth and yield in
4 different climate zones of China from 1981-2009 found that impacts were positive in Northern China
5 and negative in Southern China (Tao et al. 2014). Temperature increased across the zones while
6 precipitation changes were not consistent (Tao et al. 2014).

7 Crop yield studies focusing on India have found that warming has reduced wheat yields by 5.2% from
8 1981 to 2009, despite adaptation (Gupta et al. 2017); that maximum daytime temperatures have risen
9 along with some night-time temperatures (Jha and Tripathi 2017).

10 Agriculture in Pakistan has also been affected by climate change. From 1980 to 2014, spring maize
11 growing periods have shifted an average of 4.6 days per decade earlier, while sowing of autumn
12 maize has been delayed 3.0 days per decade (Abbas et al. 2017). A similar study with sunflower
13 showed that increases in mean temperature from 1980 to 2016 were highly correlated with shifts in
14 sowing, emergence, anthesis, and maturity for fall and spring crops (Tariq et al. 2018).

15 Mountain people in the Hindu-Kush Himalayan region encompassing parts of Pakistan, India, Nepal,
16 and China, are particularly vulnerable to food insecurity related to climate change because of poor
17 infrastructure, limited access to global markets, physical isolation, low productivity, and hazard
18 exposure, including Glacial Lake Outburst Floods (GLOFs) (Rasul et al. 2019; Rasul 2010; Tiwari
19 and Joshi 2012; Huddleston et al. 2003; Ward et al. 2013; FAO 2008; Nautiyal et al. 2007; Din et al.
20 2014). Surveys have been conducted to determine how climate-related changes have affected food
21 security (Hussain et al. 2016; Shrestha and Nepal 2016) with results showing that the region is
22 experiencing an increase in extremes, with farmers facing more frequent floods as well as prolonged
23 droughts with ensuing negative impacts on agricultural yields and increases in food insecurity
24 (Hussain et al. 2016; Manzoor et al. 2013).

25 *South America.* In another mountainous region, the Andes, inhabitants are also beginning to
26 experience changes in the timing, severity, and patterns of the annual weather cycle. Data collected
27 through participatory workshops, semi-structured interviews with agronomists, and qualitative
28 fieldwork from 2012 to 2014 suggest that in Colomi, Bolivia climate change is affecting crop yields
29 and causing farmers to alter the timing of planting, their soil management strategies, and the use and
30 spatial distribution of crop varieties (Saxena et al. 2016). In Argentina, there has also been an increase
31 in yield variability of maize and soybeans (Iizumi and Ramankutty 2016). These changes have had
32 important implications for the agriculture, human health, and biodiversity of the region (Saxena et al.
33 2016).

34 *Africa.* In recent years, yields of staple crops such as maize, wheat, sorghum, and fruit crops, such as
35 mangoes, have decreased across Africa, widening food insecurity gaps (Ketiem et al. 2017). In
36 Nigeria, there have been reports of climate change having impacts on the livelihoods of arable crop
37 farmers (Abiona et al. 2016; Ifeanyi-obi et al. 2016; Onyeneke 2018). The Sahel region of Cameroon
38 has experienced an increasing level of malnutrition, partly due to the impact of climate change since
39 harsh climatic conditions leading to extreme drought have a negative influence on agriculture
40 (Chabejong 2016).

41 Utilising farmer interviews in Abia State, Nigeria, researchers found that virtually all responders
42 agreed that the climate was changing in their area (Ifeanyi-obi et al. 2016). With regard to
43 management responses, a survey of farmers from Anambra State, Nigeria showed that farmers are
44 adapting to climate change by utilising such techniques as mixed cropping systems, crop rotation,
45 fertiliser application (Onyeneke et al. 2018). In Ebonyi State, Nigeria, Eze (2017) interviewed 160
46 women cassava farmers and found the major climate change risks in production to be severity of high
47 temperature stress, variability in relative humidity, and flood frequency.

1 *Europe*. The impacts of climate change are varied across the continent. Moore and Lobell (2015)
2 showed that climate trends are affecting European crop yields, with long-term temperature and
3 precipitation trends since 1989 reducing continent-wide wheat and barley yields by 2.5% and 3.8%,
4 respectively, and having slightly increased maize and sugar beet yields. Though these aggregate
5 affects appear small, the impacts are not evenly distributed. In cooler regions such as the United
6 Kingdom and Ireland, the effect of increased warming has been ameliorated by an increase in rainfall.
7 Warmer regions, such as Southern Europe, have suffered more from the warming; in Italy this effect
8 has been amplified by a drying, leading to yield declines of 5% or greater.

9 Another study examining the impacts of recent climate trends on cereals in Greece showed that crops
10 are clearly responding to changes in climate – and demonstrated via statistical analysis that significant
11 impacts on wheat and barley production are expected at the end of the twenty-first century
12 (Mavromatis 2015). In the Czech Republic, a study documented positive long-term impacts of recent
13 warming on yields of fruiting vegetables (cucumbers and tomatoes) (from 4.9 to 12% per 1°C
14 increase in local temperature) but decreases in yield stability of traditionally grown root vegetables in
15 the warmest areas of the country (Potopová et al. 2017). A study in Hungary also indicated the
16 increasingly negative impacts of temperature on crops and indicated that a warming climate is at least
17 partially responsible for the stagnation or reduction in crop yields since the mid-1980s in Eastern
18 Europe (Pinke and Lövei 2017).

19 In summary, climate change is already affecting some aspects of food security (*high confidence*).
20 Recent studies in both large-scale and smallholder farming systems document declines in crop
21 productivity related to rising temperatures and changes in precipitation. Evidence for climate change
22 impacts (e.g., declines and stagnation in yields, changes in sowing and harvest dates, increased
23 infestation of pests and diseases, and declining viability of some crop varieties) is emerging from
24 detection and attribution studies and indigenous and local knowledge in Australia, Europe, Asia,
25 Africa, North America, and South America (*medium evidence, robust agreement*).

26 **Projected impacts.** Climate change effects have been studied on a global scale following a variety of
27 methodologies that have recently been compared (Lobell and Asseng 2017; Zhao et al. 2017a; Liu et
28 al. 2016). Approaches to study global and local changes include global gridded crop model
29 simulations (e.g., (Deryng et al. 2014)), point-based crop model simulations (e.g., (Asseng et al.
30 2015)), analysis of point-based observations in the field (e.g., (Zhao et al. 2016)), and temperature-
31 yield regression models (e.g., (Auffhammer and Schlenker 2014)). For an evaluation of model skills
32 see e.g., used in AgMIP see Müller et al. (2017b).

33 Results from Zhao et al. (2017a) across different methods consistently showed negative temperature
34 impacts on crop yield at the global scale, generally underpinned by similar impacts at country and site
35 scales. A limitation of Zhao et al. (2017a) is that it is based on the assumption that yield responses to
36 temperature increase are linear, while yield response differs depending on growing season
37 temperature level. Iizumi et al. (2017) showed that the projected global mean yields of maize and
38 soybean at the end of this century do decrease monotonically with warming, whereas those of rice and
39 wheat increase with warming and level off at a warming of about 3°C (2091–2100 relative to 1850–
40 1900).

41 Empirical statistical models have been applied widely to different cropping systems, at multiple
42 scales. Analyses using statistical models for maize and wheat tested with global climate model
43 scenarios found that the RCP4.5 scenario reduced the size of average yield impacts, risk of major
44 slowdowns, and exposure to critical heat extremes compared to RCP8.5 in the latter decades of the
45 21st century (Tebaldi and Lobell 2018). Impacts on crops grown in the tropics are projected to be
46 more negative than in mid- to high-latitudes as stated in AR5 and confirmed by recent studies (e.g.,
47 (Levis et al. 2018)). These projected negative effects in the tropics are especially pronounced under
48 conditions of explicit nitrogen stress (Figure 5.4) (Rosenzweig et al. 2014).

GGCMs with explicit N stress



Figure 5.4 Median yield changes (%) for RCP8.5 (2070–2099 in comparison to 1980–2010 baseline) with CO₂ effects and explicit nitrogen stress over five GCMs x four Global Gridded Crop Models (GGCMs) for rainfed maize, wheat, rice, and soy (20 ensemble members from EPIC, GEPIC, pDSSAT, and PEGASUS; except for rice which has 15). Gray areas indicate historical areas with little to no yield capacity. All models use a 0.5° grid, but there are differences in grid cells simulated to represent agricultural land. While some models simulated all land areas, others simulated only potential suitable cropland area according to evolving climatic conditions; others utilised historical harvested areas in 2000 according to various data sources (Rosenzweig et al. 2014).

Reyer et al. (2017b) examined biophysical impacts in five world regions under different warming scenarios - 1, 1.5, 2, and 4 °C warming. For the Middle East and Northern Africa region a significant correlation between crop yield decrease and temperature increase was found, regardless of whether the effects of CO₂ fertilisation or adaptation measures are taken into account (Waha et al. 2017). For Latin America and the Caribbean the relationship between temperature and crop yield changes was only significant when the effect of CO₂ fertilisation is considered (Reyer et al. 2017a).

A review of recent scientific literature found that projected yield loss for West Africa depends on the degree of wetter or drier conditions and elevated CO₂ concentrations (Sultan and Gaetani 2016). Faye et al. (2018b) in a crop modelling study with RCPs 4.5 and 8.5 found that climate change could have limited effects on peanut yield in Senegal due to the effect of elevated CO₂ concentrations.

Crop productivity changes for 1.5°C and 2.0°C. The IPCC Special Report on Global Warming of 1.5°C found that climate-related risks to food security are projected to increase with global warming of 1.5°C and increase further with 2°C (IPCC 2018b). These findings are based among others on Schleussner et al. (2018); Rosenzweig et al. (2018a); Betts et al. (2018), Parkes et al. (2018) and Faye et al. (2018a). The importance of assumptions about CO₂ fertilisation was found to be significant by Ren et al. (2018) and Tebaldi and Lobell (2018)

AgMIP coordinated global and regional assessment (CGRA) results confirm that at the global scale, there are mixed results of positive and negative changes in simulated wheat and maize yields, with declines in some breadbasket regions, at both 1.5°C and 2.0°C (Rosenzweig et al. 2018a). In conjunction with price changes from the global economics models, productivity declines in the Punjab, Pakistan resulted in an increase in vulnerable households and poverty rate (Rosenzweig et al. 2018a).

Crop suitability. Another method of assessing the effects of climate change on crop yields that combined observations of current maximum-attainable yield with climate analogues also found strong reductions in attainable yields across a large fraction of current cropland by 2050 (Pugh et al. 2016).

1 However, the study found the projected total land area in 2050, including regions not currently used
2 for crops, climatically suitable for a high attainable yield similar to today. This indicates that large
3 shifts in land-use patterns and crop choice will likely be necessary to sustain production growth and
4 keep pace with current trajectories of demand.

5 *Fruits and vegetables.* Understanding the full range of climate impacts on fruits and vegetables is
6 important for projecting future food security, especially related to dietary diversity and healthy diets.
7 However, studies for vegetables are very limited (Bisbis et al. 2018). Of the 174 studies considered in
8 a recent review only 14 described results of field or greenhouse experiments studying impacts of
9 increased temperatures on yields of different root and leafy vegetables, tomatoes and legumes
10 (Scheelbeek et al. 2018). Bisbis et al. (2018) found similar effects for vegetables as have been found
11 for grain crops, that is, the effect of increased CO₂ on vegetables is mostly beneficial for production,
12 but may alter internal product quality, or result in photosynthetic down-regulation. Heat stress reduces
13 fruit set of fruiting vegetables, and speeds up development of annual vegetables, shortening their time
14 for photoassimilation. Yield losses and impaired product quality result, thereby increasing food loss
15 and waste. On the other hand, a longer growing season due to warmer temperatures enables a greater
16 number of plantings and can contribute to greater annual yields. However, some vegetables, such as
17 cauliflower and asparagus, need a period of cold accumulation to produce a harvest and warmer
18 winters may not provide those requirements.

19 For vegetables growing in higher baseline temperatures (>20°C), mean yield declines caused by 4°C
20 warming were 31.5%; for vegetables growing in cooler environments (<= 20°C), yield declines
21 caused by 4°C were much less, on the order of ~5% (Scheelbeek et al. 2018). Rippke et al. (2016)
22 found that 30–60% of the common bean growing area and 20–40% of the banana growing areas in
23 Africa will lose viability in 2078–2098 with a global temperature increase of 2.6°C and 4°C
24 respectively. Tripathi et al. (2016) found fruits and vegetable production to be highly vulnerable to
25 climate change at their reproductive stages and also due to potential for greater disease pressure.

26 In summary, studies assessed find that climate change will increasingly be detrimental to crop
27 productivity as levels of warming progress (*high confidence*). Impacts will vary depending on CO₂
28 concentrations, fertility levels, and region. Productivity of major commodity crops as well as crops
29 such as millet and sorghum yields will be affected. Studies on fruits and vegetables find similar
30 effects to those projected for grain crops in regard to temperature and CO₂ effects. Total land area
31 climatically suitable for high attainable yield, including regions not currently used for crops, will be
32 similar in 2050 to today.

34 **5.2.2.2 Impacts on livestock production systems**

35 Livestock systems are impacted by climate change mainly through increasing temperatures and
36 precipitation variation, as well as atmospheric carbon dioxide (CO₂) concentration and a combination
37 of these factors. Temperature affects most of the critical factors of livestock production, such as water
38 availability, animal production and reproduction, and animal health (mostly through heat stress)
39 (Figure 5.5). Livestock diseases are mostly affected by increases in temperature and precipitation
40 variation (Rojas-Downing et al. 2017). Impacts of climate change on livestock productivity,
41 particularly of mixed and extensive systems, are strongly linked to impacts on rangelands and
42 pastures, which include the effects of increasing CO₂ on their biomass and nutritional quality. This is
43 critical considering the very large areas concerned and the number of vulnerable people affected
44 (Steinfeld 2010; Morton 2007). Pasture quality and quantity are mainly affected through increases in
45 temperature and CO₂, and precipitation variation.

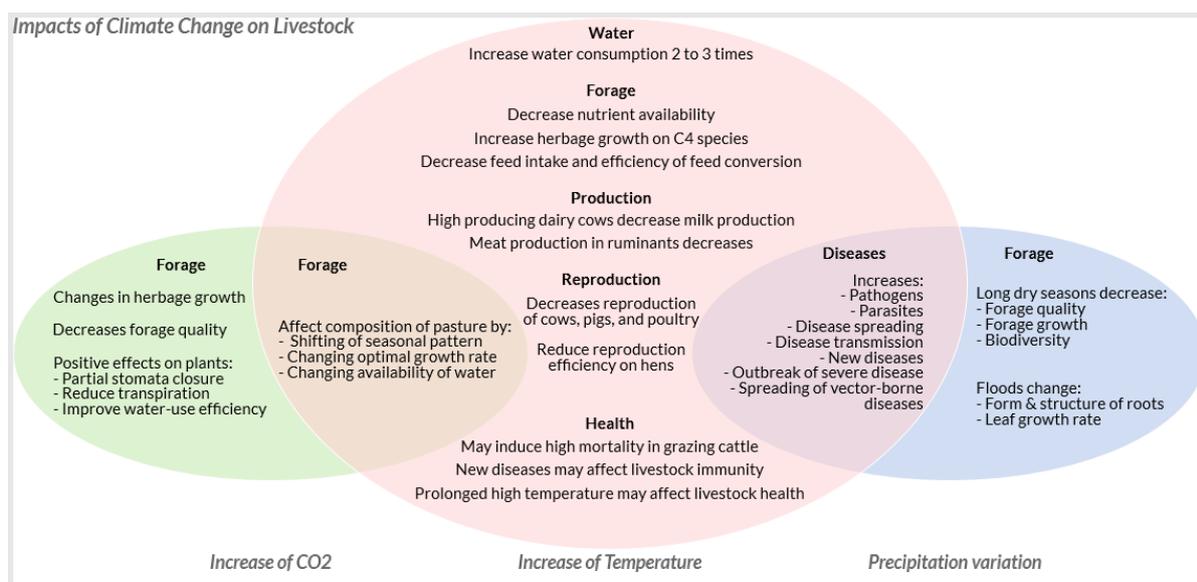


Figure 5.5 Impacts of climate change on livestock, based on (Rojas-Downing et al. 2017)

3 Among livestock systems, pastoral systems are particularly vulnerable to climate change (Dasgupta et al. 2014) (see Section 5.2.2.6 for impacts on smallholder systems that combine livestock and crops). 4 Industrial systems will suffer most from indirect impacts leading to rises in the costs of water, 5 feeding, housing, transport and the destruction of infrastructure due to extreme events, as well as an 6 increasing volatility of the price of feedstuff which increases the level of uncertainty in production 7 (Rivera-Ferre et al. 2016b; Lopez-i-Gelats 2014). Mixed systems and industrial or landless livestock 8 systems could encounter several risk factors mainly due to the variability of grain availability and 9 cost, and low adaptability of animal genotypes (Nardone et al. 2010). 10

11 Considering the diverse typologies of animal production, from grazing to industrial, Rivera-Ferre et 12 al. (2016b) distinguished impacts of climate change on livestock between those related to extreme 13 events and those related to more gradual changes in the average of climate-related variables. 14 Considering vulnerabilities, they grouped the impacts as those impacting the animal directly, such as 15 heat and cold stress, water stress, physical damage during extremes; and others impacting their 16 environment, such as modification in the geographical distribution of vector-borne diseases, location, 17 quality and quantity of feed and water and destruction of livestock farming infrastructures.

18 With severe negative impacts due to drought and high frequency of extreme events, the average gain 19 of productivity might be cancelled by the volatility induced by increasing variability in the weather. 20 For instance, semiarid and arid pasture will likely have reduced livestock productivity, while 21 nutritional quality will be affected by CO₂ fertilisation (Schmidhuber and Tubiello 2007).

22 **Observed impacts.** Pastoralism is practiced in more than 75% of countries by between 200 and 500 23 million people, including nomadic communities, transhumant herders, and agro-pastoralists 24 (McGahey et al. 2014). Observed impacts in pastoral systems reported in the literature include 25 decreasing rangelands, decreasing mobility, decreasing livestock number, poor animal health, 26 overgrazing, land degradation, decreasing productivity, decreasing access to water and feed, and 27 increasing conflicts for the access to pasture land (López-i-Gelats et al. 2016; Batima et al. 2008; 28 Njiru 2012; Fjelde and von Uexkull 2012; Raleigh and Kniveton 2012; Egeru 2016) (*high* 29 *confidence*).

30 Pastoral systems in different regions have been affected differently. For instance, in China changes in 31 precipitation were a more important factor in nomadic migration than temperature (Pei and Zhang 32 2014). There is some evidence that recent years have already seen an increase in grassland fires in

1 parts of China and tropical Asia (IPCC 2012). In Mongolia, grassland productivity has declined by
2 20-30% over the latter half of the 20th century, and ewe average weight reduced by 4 kg on an annual
3 basis, or about 8% since 1980 (Batima et al. 2008). Substantial decline in cattle herd sizes can be due
4 to increased mortality and forced off-take (Megersa et al. 2014). Important but less studied is the
5 impact of the interaction of grazing patterns with climate change on grassland composition. (Spence et
6 al. 2014) showed that climate change effects on Mongolia mountain steppe could be contingent on
7 land use.

8 Conflicts due to resource scarcity (as well as other socio-political factors (Benjaminsen et al. 2012))
9 aggravated by climate change has differentiated impact on women. In Turkana, female-headed
10 households have lower access to decision-making on resource use and allocation, investment and
11 planning (Omolo 2011), increasing their vulnerability (Section 5.1.3, Gender Box in Chapter 7).

12 Non-climate drivers add vulnerability of pastoral systems to climate change (McKune and Silva
13 2013). For instance, during environmental disasters, livestock holders have been shown to be more
14 vulnerable to food insecurity than their crop-producing counterparts because of limited economic
15 access to food and unfavorable market exchange rates (Nori et al. 2005). Sami reindeers in Finland
16 showed reduced freedom of action in response to climate change due to loss of habitat, increased
17 predation, and presence of economic and legal constraints) (Tyler et al. 2007; Pape and Löffler 2012).
18 In Tibet, emergency aid has provided shelters and privatised communally owned rangeland, which
19 have increased the vulnerability of pastoralists to climate change (Yeh et al. 2014; Næss 2013).

20 **Projected impacts.** The impacts of climate change on global rangelands and livestock have received
21 comparatively less attention than the impacts on crop production. Projected impacts on grazing
22 systems include changes in herbage growth (due to changes in atmospheric CO₂ concentrations and
23 rainfall and temperature regimes) and changes in the composition of pastures and in herbage quality,
24 as well as direct impacts on livestock (Herrero et al. 2016b). Droughts and high temperatures in
25 grasslands can also be a predisposing factor for fire occurrence (IPCC 2012).

26 *Net primary productivity, soil organic carbon, and length of growing period.* There are large
27 uncertainties related to grasslands and grazing lands (Erb et al. 2016), especially in regard to net
28 primary productivity (NPP) (Fetzel et al. 2017; Chen et al. 2018). Boone et al. (2017) estimated that
29 the mean global annual net primary production (NPP) in rangelands may decline by 10 g C m⁻² yr⁻¹ in
30 2050 under RCP 8.5, but herbaceous NPP is likely to increase slightly (i.e., average of 3 g C m⁻² yr⁻¹)
31 (Figure 5.6). Results of a similar magnitude were obtained by Havlík et al. (2015), using EPIC and
32 LPJmL on a global basis (Rojas-Downing et al. 2017). According to Rojas-Downing et al. (2017), an
33 increase of 2°C is estimate to negatively impact pasture and livestock production in arid and semiarid
34 regions and positively impact humid temperate regions.

35 Boone et al. (2017) identified significant regional heterogeneity in responses, with large increases in
36 annual productivity projected in northern regions (e.g., a 21% increase in productivity in the US and
37 Canada) and large declines in western Africa (-46% in sub-Saharan western Africa) and Australia (-
38 17%). Regarding the length of growing period (LGP, average number of growing days per year)
39 (Herrero et al. 2016b) projected reductions in the lower latitudes due to changes in rainfall patterns
40 and increases in temperatures, which indicate increasing limitations of water. They identified 35°C as
41 a critical threshold for rangeland vegetation and heat tolerance in some livestock species.



1

2

3

4

5

6

7

8

Figure 5.6 Ensemble simulation results for projected annual net primary productivity of rangelands as simulated in 2000 (top) and their change in 2050 (bottom) under emissions scenario RCP 8.5, with plant responses enhanced by CO₂ fertilisation. Results from RCP 4.5 and 8.5, with and without positive effects of atmospheric CO₂ on plant production, differed considerably in magnitude but had similar spatial patterns, and so results from RCP 8.5 with increasing production are portrayed spatially here and in other figures. Scale bar labels and the stretch applied to colors are based on the spatial mean value plus or minus two standard deviations (Boone et al. 2017).

9

10

11

12

13

14

15

16

17

18

19

Rangeland composition. According to Boone et al. (2017), the composition of rangelands is projected to change as well (see Chapter 3). Bare ground cover is projected to increase, averaging 2.4% across rangelands, with increases projected for the eastern Great Plains, eastern Australia, parts of southern Africa, and the southern Tibetan Plateau. Herbaceous cover declines are projected in the Tibetan Plateau, the eastern Great Plains, and scattered parts of the Southern Hemisphere. Shrub cover is projected to decline in eastern Australia, parts of southern Africa, the Middle East, the Tibetan Plateau, and the eastern Great Plains. Shrub cover could also increase in much of the Arctic and some parts of Africa. In mesic and semi-arid savannahs south of the Sahara, both shrub and tree cover are projected to increase, albeit at lower productivity and standing biomass. Rangelands in western and southwestern parts of the Isfahan province in Iran were found to be more vulnerable to future drying-warming conditions (Saki et al. 2018; Jaberlansar et al. 2017).

20

21

22

23

24

Soil degradation and expanding woody cover suggest that climate-vegetation-soil feedbacks catalysing shifts toward less productive, possibly stable states (Ravi et al. 2010) may threaten mesic and semi-arid savannahs south of the Sahara (see Chapter 3 and 4). This will also change their suitability for grazing different animal species; switches from cattle, which mainly consume herbaceous plants, to goats or camels are likely to occur as increases in shrubland occur.

25

26

27

Direct and indirect effects on livestock. Direct impacts of climate change in mixed and extensive production systems are linked to increased water and temperature stress on the animals potentially leading to animal morbidity, mortality and distress sales. Most livestock species have comfort zones

1 between 10°C–30°C, and at temperatures above this animals reduce their feed intake 3–5% per
2 additional degree of temperature (NRC 1981). In addition to reducing animal production, higher
3 temperatures negatively affect fertility (HLPE 2012).

4 Indirect impacts to mixed and extensive systems are mostly related to the impacts on the feed base,
5 whether pastures or crops, leading to increased variability and sometimes reductions in availability
6 and quality of the feed for the animals (Rivera-Ferre et al. 2016b). Reduced forage quality can
7 increase CH₄ emissions per unit of gross energy consumed. Increased risk of animal diseases is also
8 an important impact to all production systems (Bett et al. 2017). These depend on the geographical
9 region, land use type, disease characteristics, and animal susceptibility (Thornton et al. 2009). Also
10 important is the interaction of grazing intensity with climate change. Pfeiffer et al. (2019) estimated
11 that in a scenario of mean annual precipitation below 500 mm increasing grazing intensity reduced
12 rangeland productivity and increased annual grass abundance.

13 *Pastoral systems.* In Kenya, some 1.8 million extra cattle could be lost by 2030 because of increased
14 drought frequency, the value of the lost animals and production foregone amounting to USD 630
15 million (Herrero et al. 2010). Martin et al. (2014) assessed impacts of changing precipitation regimes
16 to identify limits of tolerance beyond which pastoral livelihoods could not be secured and found that
17 reduced mean annual precipitation had always negative effects as opposed to increased rainfall
18 variability. Similarly, Martin et al. (2016) found that drought effects on pastoralists in High Atlas in
19 Morocco depended on income needs and mobility options (see Section 5.2.2.6 for additional
20 information about impacts on smallholder farmers).

21 In summary, observed impacts in pastoral systems include changes in pasture productivity, lower
22 animal growth rates and productivity, damaged reproductive functions, increased pests and diseases,
23 and loss of biodiversity (*high confidence*). Livestock systems are projected to be adversely affected by
24 rising temperatures, depending on the extent of changes in pasture and feed quality, spread of
25 diseases, and water resource availability (*high confidence*). Impacts will differ for different livestock
26 systems and for different regions (*high confidence*). Vulnerability of pastoral systems to climate
27 change is very high (*high confidence*), and mixed systems and industrial or landless livestock systems
28 could encounter several risk factors mainly due to variability of grain availability and cost, and low
29 adaptability of animal genotypes. Pastoral system vulnerability is exacerbated by non-climate factors
30 (land tenure issues, sedentarisation programs, changes in traditional institutions, invasive species, lack
31 of markets, and conflicts) (*high confidence*).

33 **5.2.2.3 Impacts on pests and diseases**

34 Climate change is changing the dynamics of pests and diseases of both crops and livestock. The
35 nature and magnitude of future changes is likely to depend on local agro-ecological and management
36 context. This is because of the many biological and ecological mechanisms by which climate change
37 can affect the distribution, population size, and impacts of pests and diseases on food production
38 (Canto et al. 2009; Gale et al. 2009; Thomson et al. 2010; Pangga et al. 2011; Juroszek and von
39 Tiedemann 2013; Bett et al. 2017).

40 These mechanisms include changes in host susceptibility due to CO₂ concentration effects on crop
41 composition and climate stresses; changes in the biology of pests and diseases or their vectors (e.g.,
42 more generational cycles, changes in selection pressure driving evolution); mismatches in timing
43 between pests or vectors and their ‘natural enemies’; changes in survival or persistence of pests or
44 disease pathogens (e.g., changes in crop architecture driven by CO₂ fertilisation and increased
45 temperature, providing a more favourable environment for persistence of pathogens like fungi), and
46 changes in pest distributions as their “climate envelopes” shift. Such processes may affect pathogens,
47 and their vectors, as well as plant, invertebrate and vertebrate pests. (Latham et al. 2015) .

1 Furthermore, changes in diseases and their management, as well as changing habitat suitability for
2 pests and diseases in the matrix surrounding agricultural fields, have the ability to reduce or
3 exacerbate impacts (Bebber 2015). For example, changes in water storage and irrigation to adapt to
4 rainfall variation have the potential to enhance disease vector populations and disease occurrence
5 (Bett et al. 2017).

6 There is *robust evidence* that pests and diseases have already responded to climate change (Bebber et
7 al. 2014), and many studies have now built predictive models based on current incidence of pests,
8 diseases or vectors that indicate how they may respond in future (e.g., (Caminade et al. 2015; Kim et
9 al. 2015; Kim and Cho 2016; Samy and Peterson 2016; Yan et al. 2017)). Warren et al. (2018)
10 estimate that about 50% of insects, which are often pests or disease vectors, will change ranges by
11 about 50% by 2100 under current GHG emissions trajectories. These changes will lead to crop losses
12 due to changes in insect pests (Deutsch et al. 2018) and weed pressure (Ziska et al. 2018), and thus
13 affect pest and disease management at the farm level (Waryszak et al. 2018). For example, Samy and
14 Peterson (2016) modelled Blue-tongue virus (BTV), which is spread by biting *Culicoides* midges,
15 finding that the distribution of BTV is likely to be extended, particularly in central Africa, the US, and
16 western Russia.

17 There is some evidence (*medium confidence*) that exposure will, on average, increase (Bebber and
18 Gurr 2015; Yan et al. 2017), although there are a few examples where changing stresses may limit the
19 range of a vector. There is also a general expectation that perturbations may increase the likelihood of
20 pest and disease outbreaks by disturbing processes that may currently be at some quasi-equilibrium
21 (Canto et al. 2009; Thomson et al. 2010; Pangga et al. 2011). However, in some places, and for some
22 diseases, risks may decrease as well as increase (e.g., drying out may reduce the ability of fungi to
23 survive) (Kim et al. 2015; Skelsey and Newton 2015), or Tsetse fly's range may decrease (Terblanche
24 et al. 2008; Thornton et al. 2009) .

25 Pests, diseases, and vectors for both crop and livestock diseases are likely to be altered by climate
26 change (*high confidence*). Such changes are likely to depend on specifics of the local context,
27 including management, but perturbed agroecosystems are more likely, on theoretical grounds, to be
28 subject to pest and disease outbreaks (*low confidence*). Whilst specific changes in pest and disease
29 pressure will vary with geography, farming system, pest/pathogen – increasing in some situations
30 decreasing in others – there is robust evidence, with *high agreement*, that pest and disease pressures
31 are likely to change; such uncertainty requires robust strategies for pest and disease mitigation.

32 33 **5.2.2.4 Impacts on pollinators**

34 Pollinators play a key role on food security globally (Garibaldi et al. 2016). Pollinator-dependent
35 crops contribute up to 35% of global crop production volume and are important contributors to
36 healthy human diets and nutrition (IPBES 2016). On a global basis, some 1500 crops require
37 pollination (typically by insects, birds and bats) (Klein et al. 2007). Their importance to nutritional
38 security is therefore perhaps under-rated by valuation methodologies, which, nonetheless, include
39 estimates of the global value of pollination services at over USD 225 billion (2010 prices) (Hanley et
40 al. 2015). As with other ecosystem processes affected by climate change (e.g., changes in pests and
41 diseases), how complex systems respond is highly context-dependent. Thus, predicting the effects of
42 climate on pollination services is difficult (Tylianakis et al. 2008; Schweiger et al. 2010) and
43 uncertain, although there is *limited evidence* that impacts are occurring already (Section 5.2.2.4), and
44 *medium evidence* that there will be an effect.

45 Pollination services arise from a mutualistic interaction between an animal and a plant – which can be
46 disrupted by climate's impacts on one or the other or both (Memmott et al. 2007). Disruption can
47 occur through changes in species' ranges or by changes in timing of growth stages (Settele et al.

1 2016). For example, if plant development responds to different cues (e.g., day length) from insects
2 (e.g., temperature), the emergence of insects may not match the flowering times of the plants, causing
3 a reduction in pollination. Climate change will affect pollinator ranges depending on species, life-
4 history, dispersal ability and location. Warren et al. (2018) estimate that under a 3.2°C warming
5 scenario, the existing range of about 49% of insects will be reduced by half by 2100, suggesting either
6 significant range changes (if dispersal occurs) or extinctions (if it does not). However, in principle,
7 ecosystem changes caused by invasions, in some cases, could compensate for the decoupling
8 generated between native pollinators and pollinated species (Schweiger et al. 2010).

9 Other impacts include changes in distribution and virulence of pathogens affecting pollinators, such as
10 the fungus *Nosema ceranae*, which can develop at a higher temperature range than the less-virulent
11 *Nosema apis*; increased mortality of pollinators due to higher frequency of extreme weather events;
12 food shortage for pollinators due to reduction of flowering length and intensity; and aggravation of
13 other threats, such as habitat loss and fragmentation (González-Varo et al. 2013; Goulson et al. 2015;
14 Le Conte and Navajas 2008; Menzel et al. 2006; Walther et al. 2009; IPBES, 2016). The increase in
15 atmospheric CO₂ is also reducing the protein content of pollen, with potential impact on pollination
16 population biology (Ziska et al. 2016).

17 In summary, as with other complex agroecosystem processes affected by climate change (e.g.,
18 changes in pests and diseases), how pollination services respond will be highly context-dependent.
19 Thus, predicting the effects of climate on pollination services is difficult and uncertain, although there
20 is *medium evidence* that there will be an effect.

21

22 **5.2.2.5 Impacts on aquaculture**

23 This report focuses on land-based aquaculture; for assessment of impacts on marine fisheries both
24 natural and farmed see the IPCC Special Report on the Ocean and Cryosphere in a Changing Climate
25 (SROCC, forthcoming).

26 Aquaculture will be affected by both direct and indirect climate change drivers, both in the short and
27 the long-term. Barange et al. (2018) provides some examples of short-term loss of production or
28 infrastructure due to extreme events such as floods, increased risk of diseases, toxic algae and
29 parasites; and decreased productivity due to suboptimal farming conditions; and long-term impacts
30 may include scarcity of wild seed, limited access to freshwater for farming due to reduced
31 precipitation, limited access to feeds from marine and terrestrial sources, decreased productivity due
32 to suboptimal farming conditions, eutrophication and other perturbations.

33 FAO (2014a) assessed the vulnerability of aquaculture stakeholders to non-climate change drivers
34 which add to climate change hazards. Vulnerability arises from discrimination in access to inputs and
35 decision-making; conflicts; infrastructure damage; and dependence on global markets and
36 international pressures. Other non-climate drivers identified by McClanahan et al. (2015), which add
37 vulnerability to fisheries for food security include: declining fishery resources; a North–South divide
38 in investment; changing consumption patterns; increasing reliance on fishery resources for coastal
39 communities; and inescapable poverty traps creating by low net resource productivity and few
40 alternatives. In areas where vulnerability to climate change is heightened, increased exposure to
41 climate change variables and impacts is likely to exacerbate current inequalities in the societies
42 concerned, penalising further already disadvantaged groups such as migrant fishers (e.g. Lake Chad)
43 or women (e.g. employees in Chile’s processing industry) (FAO 2014a).

44 In many countries the projected declines co-occur across both marine fisheries and agricultural crops
45 (Blanchard et al. 2017), both of which will impact the aquaculture and livestock sectors (See
46 Supplementary Material Figure SM5.1). Countries with low Human Development Index, trade
47 opportunities and aquaculture technologies are likely to face greater challenges. These cross-sectoral

1 impacts point to the need for a more holistic account of the inter-connected vulnerabilities of food
2 systems to climate and global change.

3

4 **5.2.2.6 Impacts on smallholder farming systems**

5 New work has developed farming system approaches that take into account both biophysical and
6 economic processes affected by climate change and multiple activities. Farm households in the
7 developing world often rely on a complex mix of crops, livestock, aquaculture, and non-agricultural
8 activities for their livelihoods (Rosenzweig and Hillel 2015; Antle et al. 2015). Across the world,
9 smallholder farmers are considered to be disproportionately vulnerable to climate change because
10 changes in temperature, rainfall and the frequency or intensity of extreme weather events directly
11 affect their crop and animal productivity as well as their household's food security, income and well-
12 being (Vignola et al. 2015; Harvey et al. 2014b). For example, smallholder farmers in the Philippines,
13 whose survival and livelihood largely depend on the environment, constantly face risks and bear the
14 impacts of the changing climate (Peria et al. 2016).

15 Smallholder farming systems have been recognised as highly vulnerable to climate change (Morton,
16 2007) because they are highly dependent on agriculture and livestock for their livelihood (Dasgupta et
17 al. 2014) (*high confidence*). In Zimbabwe, farmers were found vulnerable due to their marginal
18 location, low levels of technology, and lack of other essential farming resources. Farmers observed
19 high frequency and severity of drought, excessive precipitation, drying up of rivers, dams and wells,
20 and changes in timing and pattern of seasons as evidence of climate change, and indicated that
21 prolonged wet, hot, and dry weather conditions resulted in crop damage, death of livestock, soil
22 erosion, bush fires, poor plant germination, pests, lower incomes, and deterioration of infrastructure
23 (Mutekwa 2009).

24 In Madagascar, Harvey et al. (2014b) conducted surveyed 600 small farmers and found that chronic
25 food insecurity, physical isolation and lack of access to formal safety nets increased Malagasy
26 farmers' vulnerability to any shocks to their agricultural system, particularly extreme events. In
27 Chitwan, Nepal, occurrence of extreme events and increased variability in temperature has increased
28 the vulnerability of crops to biotic and abiotic stresses and altered the timing of agricultural
29 operations; thereby affecting crop production (Paudel et al. 2014). In Lesotho, a study on subsistence
30 farming found that food crops were the most vulnerable to weather, followed by soil and livestock.
31 Climate variables of major concern were hail, drought and dry spells which reduced crop yields. In
32 the Peruvian Altiplan, Sietz et al. (2012) evaluate smallholders' vulnerability to weather extremes
33 with regard to food security and found the relevance of resource scarcity (livestock, land area),
34 diversification of activities (lack of alternative income, education deprivation) and income restrictions
35 (harvest failure risk) in shaping vulnerability of smallholders. See Section 5.2.2.6 for observed
36 impacts on smallholder pastoral systems.

37 **Projected impacts.** By including regional economic models, integrated methods take into account the
38 potential for yield declines to raise prices and thus livelihoods (up to a certain point) in some climate
39 change scenarios. Regional economic models of farming systems can be used to examine the potential
40 for switching to other crops and livestock, as well as the role that non-farm income can play in
41 adaptation (Valdivia et al. 2015; Antle et al. 2015). On the other hand, lost income for smallholders
42 from climate change-related declines, for example in coffee production, can decrease their food
43 security (Hannah et al. 2017).

44 Farming system methods developed by AgMIP have been used in regional integrated assessments in
45 Sub-Saharan Africa (Kihara et al. 2015), West Africa (Adiku et al. 2015); East Africa (Rao et al.
46 2015), South Africa (Beletse et al. 2015), Zimbabwe (Masikati et al. 2015), South Asia (McDermid et
47 al. 2015), Pakistan (Ahmad et al. 2015), the Indo-Gangetic Basin (Subash et al. 2015), Tamil Nadu

1 (Ponnusamy et al. 2015) and Sri Lanka (Zubair et al. 2015). The assessments found that climate
2 change adds pressure to smallholder farmers across Sub-Saharan Africa and South Asia, with winners
3 and losers within each area studied. Temperatures are expected to increase in all locations, and rainfall
4 decreases are projected for the western portion of West Africa and Southern Africa, while increases in
5 rainfall are projected for eastern West Africa and all study regions of South Asia. The studies project
6 that climate change will lead to yield decreases in most study regions except South India and areas in
7 central Kenya, as detrimental temperature effects overcome the positive effects of CO₂. These studies
8 use AgMIP representative agricultural pathways (RAPs) as a way to involve stakeholders in regional
9 planning and climate resilience (Valdivia et al. 2015). RAPs are consistent with and complement the
10 RCP/SSP approaches for use in agricultural model intercomparisons, improvement, and impact
11 assessments

12 New methods have been developed for improving analysis of climate change impacts and adaptation
13 options for the livestock component of smallholder farming systems in Zimbabwe (Descheemaeker et
14 al. 2018). These methods utilised disaggregated climate scenarios, as well as differentiating farms
15 with larger stocking rates compared to less densely stocked farms. By disaggregating climate
16 scenarios, impacts, and smallholder farmer attributes, such assessments can more effectively inform
17 decision-making towards climate change adaptation.

18 In Central Asia, a study using the bio-economic farm model (BEFM) found large differences in
19 projected climate change impact ranging from positive income gains in large-scale commercial farms
20 in contrast to negative impacts in small-scale farms (Bobojonov and Aw-Hassan 2014). Negative
21 impacts may be exacerbated if irrigation water availability declines due to climate change and
22 increased water demand in upstream regions. In Iran, changes in rainfall and water endowments are
23 projected to significantly impact crop yield and water requirements, as well as income and welfare of
24 farm families (Karimi et al. 2018).

25 Climate change impacts on food, feed and cash crops other than cereals, often grown in smallholder
26 systems or family farms are less often studied, although impacts can be substantial. For example,
27 areas suitable for growing coffee are expected to decrease by 21% in Ethiopia with global warming of
28 2.4°C (Moat et al. 2017) and more than 90% in Nicaragua (Läderach et al. 2017) with 2.2°C local
29 temperature increase.

30 Climate change can modify the relationship between crops and livestock in the landscape, affecting
31 mixed crop-livestock systems in many places. Where crop production will become marginal, livestock
32 may provide an alternative to cropping. Such transitions could occur in up to 3% of the total area of
33 Africa, largely as a result of increases in the probability of season failure in the drier mixed crop-
34 livestock systems of the continent (Thornton et al. 2014).

35 In Mexico, subsistence agriculture is expected to be the most vulnerable to climate change, due to its
36 intermittent production and reliance on maize and beans (Monterroso et al. 2014). Overall, a decrease
37 in suitability and yield is expected in Mexico and Central America for beans, coffee, maize, plantain
38 and rice (Donatti et al. 2018). Municipalities with a high proportional area under subsistence crops in
39 Central America tend to have less resources to promote innovation and action for adaptation
40 (Bouroncle et al. 2017).

41 In summary, smallholder farmers are especially vulnerable to climate change because their livelihoods
42 often depend primarily on agriculture. Further, smallholder farmers often suffer from chronic food
43 insecurity (*high confidence*). Climate change is projected to exacerbate risks of pests and diseases and
44 extreme weather events in smallholder farming systems.

45

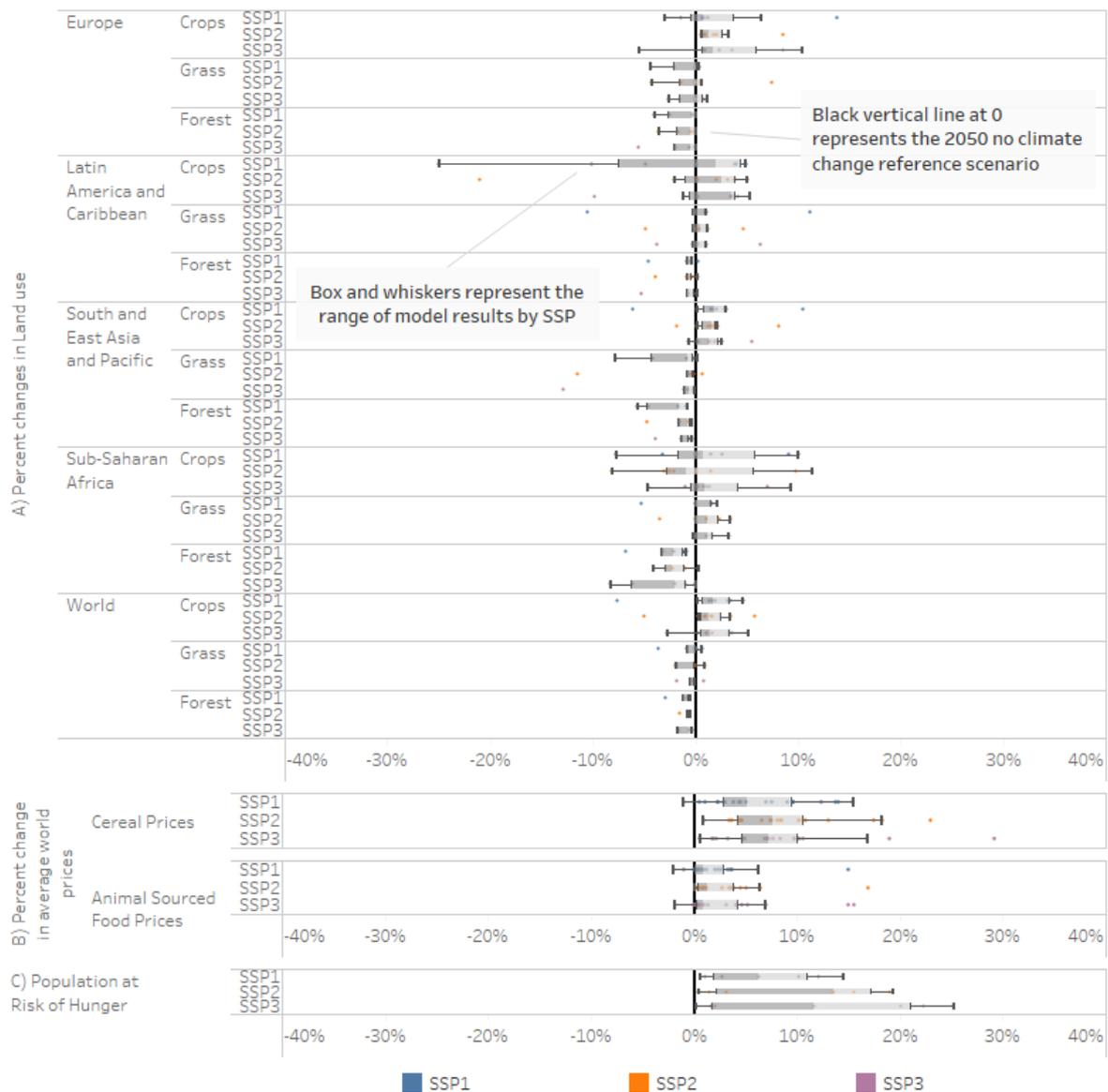
1 **5.2.3 Climate change impacts on access**

2 Access to food involves the ability to obtain food, including the ability to purchase food at affordable
3 prices.

4

5 **5.2.3.1 Impacts on prices and risk of hunger**

6 A protocol-based analysis based on AgMIP methods tested a combination of RCPs and SSPs to
7 provide a range of projections for prices, risk of hunger, and land use change (Figure 5.7 and
8 Supplementary Material Table SM5.4.) (Hasegawa et al. 2018). Previous studies have found that
9 decreased agricultural productivity will depress agricultural supply, leading to price increases. Despite
10 different economic models with various representations of the global food system (Valin et al. 2014;
11 Robinson et al. 2014; Nelson et al. 2013; Schmitz et al. 2014), as well as having represented the SSPs
12 in different ways (i.e., technological change, land-use policies, sustainable diets, etc. (Stehfest et al.
13 2019; Hasegawa et al. 2018)), the ensemble of participating models projected a 1-29% cereal price
14 increase in 2050 across SSPs 1, 2, and 3 due to climate change (RCP 6.0), which would impact
15 consumers globally through higher food prices; regional effects will vary. The median cereal price
16 increase was 7%, given current projections of demand. In all cases (across SSPs and global economic
17 models), prices are projected to increase for rice and coarse grains, with only one instance of a price
18 decline (-1%) observed for wheat in SSP1, with price increases projected in all other cases. Animal-
19 sourced foods (ASFs) are also projected to see price increases (1%), but the range of projected price
20 changes are about half those of cereals, highlighting that the climate impacts on ASFs is indirect
21 through the cost and availability of feed, and that there is significant scope for feed substitution within
22 the livestock sector.



1
 2 **Figure 5.7. Implications of climate change by 2050 on land-use, selected agricultural commodity prices,**
 3 **and the population at risk of hunger based on AgMIP Global Economic Model analysis. (A) Projected %**
 4 **change in land-use by 2050 by land type (cropland, grassland, and forest) and SSP. (B) Projected %**
 5 **changes in average world prices by 2050 for cereals (rice, wheat, and coarse grains) and animal sourced**
 6 **foods (ruminant meat, monogastric, and dairy) by SSP. (C) Percentage change by 2050 in the global**
 7 **population at risk of hunger by SSP.**

8 Declining food availability caused by climate change is likely to lead to increasing food cost
 9 impacting consumers globally through higher prices and reduced purchasing power, with low-income
 10 consumers particularly at risk from higher food prices (Nelson et al. 2010; Springmann et al. 2016a;
 11 Nelson et al. 2018). Higher prices depress consumer demand, which in turn will not only reduce
 12 energy intake (calories) globally (Hasegawa et al. 2015; Nelson et al. 2010; Springmann et al. 2016a;
 13 Hasegawa et al. 2018), but will also likely lead to less healthy diets with lower availability of key
 14 micronutrients (Nelson et al. 2018) and increase diet-related mortality in lower and middle-income
 15 countries (Springmann et al. 2016a). These changes will slow progress towards the eradication of
 16 malnutrition in all its forms.

17 The extent that reduced energy intake leads to a heightened risk of hunger varies by global economic
 18 model. However, all models project an increase in the risk of hunger, with the median projection of an

1 increase in the population at risk of insufficient energy intake by 6, 14, and 12 % in 2050 for SSPs 1,
2 2 and 3 respectively compared to a no climate change reference scenario. This median percentage
3 increase would be the equivalent of 8, 24, and 80 million (full range 1-183 million) additional people
4 at risk of hunger due to climate change (Hasegawa et al. 2018).

6 **5.2.3.2 Impacts on land use**

7 Climate change is likely to lead to changes in land use globally (Nelson et al. 2014; Schmitz et al.
8 2014; Wiebe et al. 2015). Hasegawa et al. (2018) found that declining agricultural productivity
9 broadly leads to the need for additional cropland, with 7 of 8 models projecting increasing cropland
10 and the median increase by 2050 projected across all models of 2 % compared to a no climate change
11 reference (Figure 5.7). Not all regions will respond to climate impacts equally, with more uncertainty
12 on regional land-use change across the model ensemble than the global totals might suggest. For
13 example, the median land-use change for Latin America is an increase of cropland by 3 %, but the
14 range across the model ensemble is significant, with 3 models projecting declines in cropland (-25 –-1
15 %) compared to the 5 models projecting cropland increase (0 – 5 %). For further discussion on land
16 use change and food security see Section 5.6.

18 **5.2.4 Climate change impacts on food utilisation**

19 Food utilisation involves nutrient composition of food, its preparation, and overall state of health.
20 Food safety and quality affects food utilisation.

22 **5.2.4.1 Impacts on food safety and human health**

23 Climate change can influence food safety through changing the population dynamics of contaminating
24 organisms due to, for example, changes in temperature and precipitation patterns, and also humidity,
25 increased frequency and intensity of extreme weather events, and changes in contaminant transport
26 pathways. Changes in food and farming systems, e.g., intensification to maintain supply under climate
27 change, may also increase vulnerabilities as the climate changes (Tirado et al. 2010). Climate-related
28 changes in the biology of contaminating organisms include changing the activity of mycotoxin-
29 producing fungi, changing the activity of micro-organisms in aquatic food chains that cause disease
30 (e.g., dinoflagellates, bacteria like *Vibrio*), and increasingly heavy rainfall and floods causing
31 contamination of pastures with enteric microbes (like *Salmonella*) that can enter the human food
32 chain. Degradation and spoilage of products in storage and transport can also be affected by changing
33 humidity and temperature outside of cold chains, notably from microbial decay but also from potential
34 changes in the population dynamics of stored product pests (e.g., mites, beetles, moths) (Moses et al.
35 2015).

36 Mycotoxin-producing fungi occur in specific conditions of temperature and humidity, so climate
37 change will affect their range, increasing risks in some areas (such as mid-temperate latitudes) and
38 reducing them in others (e.g., the tropics) (Paterson and Lima 2010). There is *robust evidence* from
39 process-based models of particular species (*Aspergillus*/Aflatoxin B1, *Fusarium*/deoxynivalenol) with
40 projections of future climate that show that aflatoxin contamination of maize in southern Europe will
41 increase significantly (Battilani et al. 2016), and deoxynivalenol contamination of wheat in north-west
42 Europe will increase by up to 3 times (van der Fels-Klerx et al. 2012b,a). Whilst the downscaled
43 climate models make any specific projection for a given geography uncertain (Van der Fels-Klerx et
44 al. 2013), experimental evidence on the small scale suggests that the combination of rising CO₂ levels,
45 affecting physiological processes in photosynthetic organisms, and temperature changes, can be
46 significantly greater than temperature alone (Medina et al. 2014). Risks related to aflatoxins are likely

1 to change, but detailed projections are difficult because they depend on local conditions (Vaughan et
2 al. 2016).

3 Foodborne pathogens in the terrestrial environment typically come from enteric contamination (from
4 humans or animals), and can be spread by wind (blowing contaminated soil) or flooding – the
5 incidence of both of which are likely to increase with climate change (Hellberg and Chu 2016).
6 Furthermore, water stored for irrigation, which may be increased in some regions as an adaptation
7 strategy, can become an important route for the spread of pathogens (as well as other pollutants);
8 contaminated water and diarrheal diseases are acute threats to food security (Bond et al. 2018). Whilst
9 there is little direct evidence (in terms of modelled projections) the results of a range of reviews, as
10 well as expert groups, suggest that risks from foodborne pathogens are likely to increase through
11 multiple mechanisms (Tirado et al. 2010; van der Spiegel et al. 2012; Liu et al. 2013; Kirezieva et al.
12 2015; Hellberg and Chu 2016).

13 An additional route to climate change impacts on human health can arise from the changing biology
14 of plants altering human exposure levels. This may include climate changing how crops sequester
15 heavy metals (Rajkumar et al. 2013), or how they respond to changing pest pressure (e.g., cassava
16 produces hydrogen cyanide as a defence against herbivore attack).

17 All of these factors will lead to regional differences regarding food safety impacts (Paterson and Lima
18 2011). For instance, in Europe it is expected that most important food safety-related impacts will be
19 mycotoxins formed on plant products in the field or during storage; residues of pesticides in plant
20 products affected by changes in pest pressure; trace elements and/or heavy metals in plant products
21 depending on changes in abundance and availability in soils; polycyclic aromatic hydrocarbons in
22 foods following changes in long-range atmospheric transport and deposition; and presence of
23 pathogenic bacteria in foods following more frequent extreme weather, such as flooding and heat
24 waves (Miraglia et al. 2009).

25 In summary, there is *medium evidence*, with *high agreement* that food utilisation via changes in food
26 safety (and potentially food access from food loss) will be impacted by climate change, mostly by
27 increasing risks, but there is *low confidence*, exactly how they may change for any given place.

28

29 **5.2.4.2 Impacts on food quality**

30 There are two main routes by which food quality may change. First, the direct effects of climate
31 change on plant and animal biology, such as through changing temperatures changing the basic
32 metabolism of plants. Secondly, by increasing carbon dioxide's effect on biology through CO₂
33 fertilisation.

34 *Direct effects on plant and animal biology.* Climate affects a range of biological processes, including
35 the metabolic rate in plants and ectothermic animals. Changing these processes can change growth
36 rates, and therefore yields, but can also cause organisms to change relative investments in growth vs
37 reproduction, and therefore change the nutrients assimilated. This may decrease protein and mineral
38 nutrient concentrations, as well as alter lipid composition (DaMatta et al. 2010). For example, apples
39 in Japan have been exposed to higher temperatures over 3–4 decades and have responded by
40 blooming earlier. This has led to changes in acidity, firmness, and water content, reducing quality
41 (Sugiura et al. 2013). In other fruit, such as grapes, warming-induced changes in sugar composition
42 affect both colour and aroma (Mira de Orduña 2010). Changing heat stress in poultry can affect yield
43 as well as meat quality (by altering fat deposition and chemical constituents), shell quality of eggs,
44 and immune systems (Lara and Rostagno 2013).

45 *Effects of rising CO₂ concentrations.* Climate change is being driven by rising concentrations of
46 carbon dioxide and other greenhouse gases in the atmosphere. As plants use CO₂ in photosynthesis to

1 form sugar, rising CO₂ levels, all things being equal, enhances the process unless limited by water or
2 nitrogen availability. This is known as “CO₂ fertilisation”. Furthermore, increasing CO₂ allows the
3 stomata to be open for a shorter period for gas exchange, reducing water loss through transpiration.
4 These two factors affect the metabolism of plants, and, as with changing temperatures, affects plant
5 growth rates, yields and their nutritional quality. Studies of these effects include meta-analyses,
6 modelling, and small-scale experiments (Franzaring et al. 2013; Mishra and Agrawal 2014; Myers et
7 al. 2014; Ishigooka et al. 2017; Zhu et al. 2018; Loladze 2014; Yu et al. 2014)

8 In regard to nutrient quality, a meta-analysis from seven Free-Air Carbon dioxide Enrichment
9 (FACE), (with elevated atmospheric CO₂ concentration of 546–586 ppm) experiments (Myers et al.
10 2014), found that wheat grains had 9.3% lower zinc (CI 5.9–12.7%), 5.1% lower iron (CI 3.7–6.5%)
11 and 6.3% lower protein (CI 5.2–7.5%), and rice grains had 7.8% lower protein content (CI 6.8–8.9%).
12 Changes in nutrient concentration in field pea, soybean and C4 crops such as sorghum and maize were
13 small or insignificant. Zhu et al. (2018) report a meta-analysis of FACE trials on a range of rice
14 cultivars. They show that protein declines by an average of 10% under elevated CO₂, iron and zinc
15 decline by 8% and 5% respectively. Furthermore, a range of vitamins show large declines across all
16 rice cultivars, including B1 (-17%), B2 (-17%), B5 (-13%) and B9 (-30%), whereas Vitamin E
17 increased. As rice underpins the diets of many of the world’s poorest people in low-income countries,
18 especially in Asia, Zhu et al. (2018) estimate that these changes under high CO₂ may affect the
19 nutrient status of about 600 million people.

20 Decreases in protein concentration with elevated CO₂ are related to reduced nitrogen concentration
21 possibly caused by nitrogen uptake not keeping up with biomass growth, an effect called
22 ‘carbohydrate dilution’ or ‘growth dilution’, and by inhibition of photorespiration which can provide
23 much of the energy used for assimilating nitrate into proteins (Bahrami et al. 2017). Other
24 mechanisms have also been postulated (Feng et al. 2015; Bloom et al. 2014; Taub and Wang 2008).
25 Together, the impacts on protein availability may take as many as 150 million people into protein
26 deficiency by 2050 (Medek et al. 2017). Legume and vegetable yields increased with elevated CO₂
27 concentration of 250 ppm above ambient by 22% (CI 11.6–32.5%), with a stronger effect on leafy
28 vegetables than on legumes and no impact for changes in iron, vitamin C or flavonoid concentration
29 (Scheelbeek et al. 2018).

30 Increasing concentrations of atmospheric CO₂ lower the content of zinc and other nutrients in
31 important food crops. Dietary deficiencies of zinc and iron are a substantial global public health
32 problem (Myers et al. 2014). An estimated two billion people suffer these deficiencies (FAO 2013a),
33 causing a loss of 63 million life-years annually (Myers et al. 2014). Most of these people depend on
34 C3 grain legumes as their primary dietary source of zinc and iron. Zinc deficiency is currently
35 responsible for large burdens of disease globally, and the populations who are at highest risk of zinc
36 deficiency receive most of their dietary zinc from crops (Myers et al. 2015). The total number of
37 people estimated to be placed at new risk of zinc deficiency by 2050 is 138 million. The people likely
38 to be most affected live in Africa and South Asia, with nearly 48 million residing in India alone.
39 Differences between cultivars of a single crop suggest that breeding for decreased sensitivity to
40 atmospheric CO₂ concentration could partly address these new challenges to global health (Myers et
41 al. 2014).

42 In summary, while increased CO₂ is projected to be beneficial for crop productivity at lower
43 temperature increases, it is projected to lower nutritional quality (e.g., less protein, zinc, and iron)
44 (*high confidence*).

45

1 **5.2.5 Climate change impacts on food stability**

2 Food stability is related to people's ability to access and use food in a steady way, so that there not
3 intervening periods of hunger. Increasing extreme events associated with climate change can disrupt
4 food stability. (See Section 5.8.1 for assessment of food price spikes.)

5

6 **5.2.5.1 Impacts of extreme events**

7 FAO et al. (2018) conducted an analysis of prevalence of undernourishment (PoU) and found that in
8 2017, the average of the prevalence of undernourishment (PoU) was 15.4% for all countries exposed
9 to climate extremes (See Supplementary Material Figure SM5.2). At the same time, the PoU was 20%
10 for countries that additionally show high vulnerability of agriculture production/yields to climate
11 variability, or 22.4% for countries with high PoU vulnerability to severe drought. When there is both
12 high vulnerability of agriculture production/yields and high PoU sensitivity to severe drought, the
13 PoU is 9.8 points higher (25.2%). These vulnerabilities were found to be higher when countries had a
14 high dependence on agriculture as measured by the number of people employed in the sector.
15 Bangkok experienced severe flooding in 2011-2012 with large-scale disruption of the national food
16 supply chains since they were centrally organised in the capital city (Allen et al. 2017).

17 The IPCC projects that frequency, duration, and intensity of some extreme events will increase in the
18 coming decades (IPCC 2018a, 2012). To test these effects on food security, Tigchelaar et al. (2018)
19 showed rising instability in global grain trade and international grain prices, affecting especially the
20 about 800 million people living in extreme poverty who are most vulnerable to food price spikes (see
21 Section 5.8.1). They used global datasets of maize production and climate variability combined with
22 future temperature projections to quantify how yield variability will change in the world's major
23 maize-producing and -exporting countries under 2°C and 4°C of global warming.

24 Tesfaye et al. (2017) projected that the extent of heat-stressed areas in South Asia could increase by
25 up to 12% in 2030 and 21% in 2050 relative to the baseline (1950–2000). Another recent study found
26 that drier regions are projected to dry earlier, more severely and to a greater extent than humid
27 regions, with the population of sub-Saharan Africa most vulnerable (Lickley and Solomon 2018).

28

29 **5.2.5.2 Food aid**

30 Food aid plays an important role in providing food security and saving lives after climate disasters. In
31 2015, 14.5 million people were assisted through disaster-risk reduction, climate change and/or
32 resilience building activities (WFP 2018). However, there is no agreement on how to better use
33 emergency food aid, since it can come with unintended consequences for individuals, groups, regions,
34 and countries (Barrett 2006). These may include negative dependency of food recipients (Lentz et al.
35 2005) or price increases, among others.

36 Some authors state that tied food aid provided as “in kind” by the donor country hampers local food
37 production (Clay 2006), although others found no evidence of this (Ferrière and Suwa-Eisenmann
38 2015). Untied cash aid can be used to buy food locally or in neighbouring countries, which is cheaper
39 and can contribute to improving the livelihoods of local farmers (Clay 2006).

40 Ahlgren et al. (2014) found that food aid dependence of Marshall Islands due to climate change
41 impacts can result in poor health outcomes due to the poor nutritional quality of food aid, which may
42 result in future increases of chronic diseases. In this regard, Mary et al. (2018) showed that nutrition-
43 sensitive aid can reduce the prevalence of undernourishment.

44 In summary, based on AR5 and SR15 assessments that the likelihood that extreme weather will
45 increase, (e.g., increases in heatwaves, droughts, inland and coastal flooding due to sea level rise

1 depending on region) in both frequency and magnitude, decreases in food stability and thus increases
2 in food insecurity will likely rise as well (*medium evidence, high agreement*).

4 **5.3 Adaptation options, challenges, and opportunities**

5 This section assesses the large body of literature on food system adaptation to climate change,
6 including increasing extreme events, within a framework of autonomous, incremental, and
7 transformational adaptation. It focuses primarily on regional and local considerations and adaptation
8 options for both the supply side (production, storage, transport, processing, and trade) and the demand
9 side (consumption and diets) of the food system. Agroecological, social, and cultural contexts are
10 considered throughout. Finally, the section assesses the role of institutional measures at global,
11 regional (multiple countries), national, and local scales and capacity-building.

13 **5.3.1 Challenges and opportunities**

14 By formulating effective adaptation strategies, it is possible to reduce or even avoid some of the
15 negative impacts of climate change on food security (See Section 5.2). However, if unabated climate
16 change continues, limits to adaptation will be reached (SR15). In the food system, adaptation actions
17 involve any activities designed to reduce vulnerability and enhance resilience of the system to climate
18 change. In some areas, expanded climate envelopes will alter agro-ecological zones, with opportunity
19 for expansion towards higher latitudes and altitudes, soil and water resources permitting (Rosenzweig
20 and Hillel 2015).

21 More extreme climatic events are projected to lead to more agro-meteorological disasters with
22 associated economic and social losses. There are many options for adapting the food system to
23 extreme events reported in IPCC (2012), highlighting measures that reduce exposure and vulnerability
24 and increase resilience, even though risks cannot fully be eliminated (IPCC 2012). Adaptation
25 responses to extreme events aim to minimise damages, modify threats, prevent adverse impacts, or
26 share losses, thus making the system more resilient (Harvey et al. 2014a).

27 With current and projected climate change (higher temperature, changes in precipitation, flooding and
28 extremes events), achieving adaptation will require both technological (e.g., recovering and improving
29 orphan crops, new cultivars from breeding or biotechnology) and non-technological (e.g., market,
30 land management, diet change) solutions. Climate interacts with other factors such as market supplies
31 over longer distances and policy drivers (Mbow et al. 2008; Howden et al. 2007), as well as local
32 agricultural productivity to determine access to food locally.

33 Given the site-specific nature of climate change impacts on food system components together with
34 wide variation in agroecosystems types and management, and socio-economic conditions, it is widely
35 understood that adaptation strategies are linked to environmental and cultural contexts at the regional
36 and local levels (*high confidence*). Developing systemic resilience that integrates climate drivers with
37 social and economic drivers would reduce the impact on food security, particularly in developing
38 countries. For example, in Africa, improving food security requires evolving food systems to be
39 highly climate resilient, while supporting the need for increasing yield to feed the growing population
40 (Mbow et al. 2014b) (Box 5.2).

41 Adaptation involves producing more food where needed, moderating demand, reducing waste, and
42 improving governance (Godfray and Garnett 2014) (see Section 5.6 for the significant synergies
43 between adaptation and mitigation through specific practices, actions and strategies.).

1 **Box 5.2 Sustainable solutions for food systems and climate change in Africa**

2 Climate change, land use change, and food security are important aspects of sustainability policies in
3 Africa. According to the McKinsey Global Institute (2010), Africa has around 60% of the global
4 uncultivated arable land; thus the continent has a high potential for transformative change in food
5 production. With short and long-term climate change impacts combined with local poverty conditions,
6 land degradation and poor farming practices, Africa cannot grow enough food to feed its rapidly
7 growing population. Sustainable improvement of productivity is essential, even as the impacts of
8 climate change on food security in Africa are projected to be plural and severe.

9 Sustainable Land Management (SLM) of farming systems is important to address climate change
10 while dealing with these daunting food security needs and the necessity to improve access to
11 nutritious food to maintain healthy and active lives in Africa (AGRA 2017). SLM has functions
12 beyond the production of food, such as delivery of water, protection against disease (especially
13 zoonotic diseases), the delivery of energy, fibre and building materials.

14 Commodity-based systems—driven by external markets—are increasing in Africa (cotton, cocoa,
15 coffee, oil palm, groundnuts) with important impacts on the use of land and climate. Land
16 degradation, decreasing water resources, loss of biodiversity, excessive use of synthetic fertilisers and
17 pesticides are some of the environmental challenges that influence preparedness to adapt to climate
18 change (Pretty and Bharucha 2015).

19 A balanced strategy on African agriculture can be based on SLM and multifunctional land use
20 approaches combining food production, cash crops, ecosystem services, biodiversity conservation,
21 and ecosystem services delivery, and indigenous and local knowledge.

22 Thus, sustainable food systems in Africa entail multiple dimensions as shown in Figure 5.7.



24
25 **Figure 5.7 Factors influencing sustainable food systems in Africa**

26 With rapid urbanisation, it is important to use combined land goals (e.g., zero-carbon energy, smart
27 irrigation systems, and climate-resilient agriculture) to minimise the negative side effects of climate
28 change while securing quality food for a growing population.

29 Building resilience into productivity and production can be based on simultaneous attention to the
30 following five overarching issues:

31 1) Closing yield gaps through adapted cultivars, sustainable land management, that
32 combine production and preservation of ecosystems essential functions such as sustainable

1 intensification approaches based on conservation agriculture and community-based adaptation with
2 functioning support services and market access (Mbow et al. 2014a).

3 2) Identifying Sustainable Land Management practices (agroecology, agroforestry, etc.)
4 addressing different ecosystem services (food production, biodiversity, reduction of GHG emissions,
5 soil carbon sequestration) for improved land-based climate change adaptation and mitigation (Sanz et
6 al. 2017; Francis 2016).

7 3) Paying attention to the food-energy-water nexus, especially water use and
8 reutilisation efficiency but also management of rain water (Albrecht et al. 2018).

9 4) Implementing institutional designs focused on youth, women through new economic
10 models that help access credit and loans to support policies that balance cash and food crops.

11 5) Build on and use of local knowledge, culture and traditions while seeking innovations
12 for food waste reduction and transformation of agricultural products.

13 These aspects suppose both incremental and transformational adaptation that may stem from better
14 infrastructure (storage and food processing), adoption of harvest and post-harvest technologies that
15 minimise food waste, and development of new opportunities for farmers to respond to environmental,
16 economic and social shocks that affect their livelihoods (Morton 2017).

17 Agriculture in Africa offers a unique opportunity for merging adaptation to and mitigation of climate
18 change with sustainable production to ensure food security (CCAFS 2012; FAO 2012). Initiatives
19 throughout the food system on both the supply and demand sides can lead to positive outcomes.

21 **5.3.2 Adaptation framing and key concepts**

22 **5.3.2.1 Autonomous, incremental, and transformational adaptation**

23 Framing of adaptation in this section categorises and assesses adaptation measures as autonomous,
24 incremental, and transformational (See Glossary and Table 5.3). Adaptation responses can be reactive
25 or anticipatory.

26 *Autonomous.* Autonomous adaptation in food systems does not constitute a conscious response to
27 climatic stimuli but is triggered by changes in agroecosystems, markets, or welfare changes. It is also
28 referred to as spontaneous adaptation (IPCC 2007). Examples of autonomous adaptation of rural
29 populations have been documented in the Sahel (IRD 2017). In India, farmers are changing sowing
30 and harvesting timing, cultivating short duration varieties, inter-cropping, changing cropping patterns,
31 investing in irrigation, and establishing agroforestry. These are considered as passive responses or
32 autonomous adaptation, because they do not acknowledge that these steps are taken in response to
33 perceived climatic changes (Tripathi and Mishra 2017).

34 *Incremental.* Incremental adaptation maintains the essence and integrity of a system or process at a
35 given scale (Park et al. 2012). Incremental adaptation focuses on improvements to existing resources
36 and management practices. The central aim of incremental adaptation is to maintain the essence and
37 integrity of a system or process at a given scale (IPCC 2014a).

38 *Transformational.* Transformational adaptation changes the fundamental attributes of a socio-
39 ecological system either in anticipation of or in response to climate change and its impacts (IPCC
40 2014a). Transformational adaptation seeks alternative livelihoods and land use strategies needed to
41 develop new farming systems (Termeer et al. 2016). For example, limitations in incremental
42 adaptation among smallholder rice farmers in Northwest Costa Rica led to a shift from rice to
43 sugarcane production due to decreasing market access and water scarcity (Warner et al. 2015).
44 Migration from the Oldman River Basin has been described as a transformational adaptation to climate

1 change in the Canadian agriculture sector (Hadarits et al. 2017). If high-end scenarios of climate
 2 change eventuate, the food security of farmers and consumers will depend on how transformational
 3 change in food systems is managed. An integrated framework of adaptive transition – management of
 4 socio-technical transitions and adaptation to socio-ecological changes – may help build
 5 transformational adaptive capacity (Mockshell and Kamanda 2018; Pant et al. 2015). Rippke et al.
 6 (2016) has suggested overlapping phases of adaptation needed to support transformational change in
 7 Africa.

8
 9 **Table 5.3 Synthesis of food security related adaptation options to address various climate risks (IPCC**
 10 **2014b; Vermeulen et al. 2013, 2018; Burnham and Ma 2016; Bhatta and Aggarwal 2016)**

| Key climate drivers and risks | Incremental adaptation | Transformational adaptation | Enabling conditions |
|---|--|--|--|
| <p>Extreme events and short-term climate variability</p> <p>Stress on water resources, drought stress, dry spells, heat extremes, flooding, shorter rainy seasons, pests</p> | <ul style="list-style-type: none"> - Change in variety, water management, water harvesting, supplemental irrigation during dry spells, - Planting dates, pest control, feed banks, - Transhumance, Other sources of revenue (e.g. charcoal, wild fruits, wood, temporary work) - Soil management, composting, | <ul style="list-style-type: none"> - Early Warning Systems - Use of planning and prediction at seasonal to intra-seasonal climate risk to transition to a food safer condition. - Abandonment of monoculture, diversification - Crop and livestock insurance - Alternate cropping, intercropping -Erosion control | <ul style="list-style-type: none"> - Establishment of climate services - Integrated water management policies, integrated land and water governance - Seed banks, seed sovereignty and seed distribution policies - Capacity building and extension programs |
| <p>Warming trend, drying trend</p> <p>Reduced crop productivity due to persistent heat, long drought cycles, deforestation and land degradation with strong adverse effects on food production and nutrition quality, increased pest and disease damage</p> | <ul style="list-style-type: none"> - Strategies to reduce effects of recurring food challenges - Sustainable intensification, agroforestry, conservation agriculture, SLM - Adoption of existing drought-tolerant crop and livestock species - Counter season crop production, - Livestock fattening - New ecosystem-based adaptation (e.g. bee keeping, woodlots) | <ul style="list-style-type: none"> - Climate services for new agricultural programs, e.g., sustainable irrigation districts) - New technology, e.g., new farming systems, new crops and livestock breeds - Switches between cropping and transhumant livelihoods, replacement of pasture or forest to irrigated/rainfed crops - Shifting to small ruminants or drought resistant livestock or fish | <ul style="list-style-type: none"> - Climate information in local development policies. - Stallholders' access to credit and production resources, - National food security program based on increased productivity, diversification, transformation and trade - Strengthening (budget, capacities, expertise) of local and national institutions to support agriculture and |

| | | | |
|--|---|---|--|
| | <ul style="list-style-type: none"> - Farmers management of natural resources - Labor redistribution (e.g., mining, development projects, urban migration) - Adjustments to markets and trade pathways already in place | <ul style="list-style-type: none"> farming Food storage infrastructures, food transformation - Changes in cropping area, land rehabilitation (enclosures, afforestation) perennial farming - New markets and trade pathways | <ul style="list-style-type: none"> livestock breeding - Devolution to local communities, women empowerment, market opportunities - Incentives for establishing new markets and trade pathways |
|--|---|---|--|

1

2 **5.3.2.2 Risk management**

3 Climate risks affect all pillars of food security, particularly stability because extreme events lead to
4 strong variation to food access. The notion of risk is widely treated in IPCC reports (IPCC 2014c) (see
5 also Chapter 7 in this report). With food systems, many risks co-occur or reinforce each other and this
6 can limit effective adaptation planning as they require a comprehensive and dynamic policy approach
7 covering a range of drivers and scales. For example, from the understanding by farmers of change in
8 risk profiles to the establishment of efficient markets that facilitate response strategies will require
9 more than systemic reviews of risk factors (Howden et al. 2007).

10 Integration of Climate Change Adaptation (CCA) and Disaster Risk Reduction (DRR) helps to
11 minimise the overlap and duplication of projects and programs (Nalau et al. 2016). Recently,
12 countries started integrating the concept of DRR and CCA. For instance, The Philippines has
13 introduced legislation calling for CCA and DRR integration as current policy instruments were largely
14 unsuccessful in combining agencies and experts across the two areas (Leon and Pittock 2016).

15 Studies reveal that the amplitude of interannual growing-season temperature variability is in general
16 larger than that of long-term temperature change in many locations. Responding better to seasonal
17 climate-induced food supply shocks therefore increases society's capability to adapt to climate
18 change. Given these backgrounds, seasonal crop forecasting and early response recommendations,
19 based on seasonal climate forecasts, are emerging to strengthen existing operational systems for
20 agricultural monitoring and forecasting (FAO 2016a; Ceglar et al. 2018; Iizumi et al. 2018).

21 While adaptation and mitigation measures are intended to reduce the risk from climate change
22 impacts in food systems, they can also be sources of risk themselves (e.g. investment risk, political
23 risk) (IPCC 2014b). Climate-related hazards are a necessary element of risks related to climate
24 impacts but may have little or nothing to do with risks related to some climate policies/responses

25 Adoption of agroecological practices could provide resilience for future shocks, spread farmer risk
26 and mitigate the impact of droughts (Niles et al. 2018) (See Section 5.3.2.3). Traditionally, risk
27 management is performed through multifunctional landscape approaches in which resource utilisation
28 is planned across wide areas and local agreements on resource access. Multifunctionality permits
29 vulnerable communities to access various resources at various times and under various risk conditions
30 (Minang et al. 2015).

31 In many countries, governmental compensation for crop-failure and financial losses are used to
32 protect against risk of severe yield reductions. Both public and private sector groups develop
33 insurance markets and improve and disseminate index-based weather insurance programs.

1 Catastrophe bonds, microfinance, disaster contingency funds, and cash transfers are other available
2 mechanisms for risk management.

3 In summary, risk management can be accomplished through agro-ecological landscape approaches
4 and risk sharing and transfer mechanisms such as development of insurance markets and improve
5 index-based weather insurance programs (*high confidence*).

6

7 **5.3.2.3 Role of agroecology and diversification**

8 Agro-ecological systems are integrated land-use systems that maintain species diversity in a range of
9 productive niches. Diversified cropping systems and practicing traditional agro-ecosystems of crop
10 production where a wide range of crop varieties are grown in various spatial and temporal
11 arrangements, are less vulnerable to catastrophic loss (Zhu et al. 2011). The use of local genetic
12 diversity, soil organic matter enhancement, multiple-cropping or poly-culture systems, and home
13 gardening, agro-ecological approaches can build resilience against extreme climate events (Altieri and
14 Koohafkan 2008). However, Nie et al. (2016) argued that while integrated crop-livestock systems
15 present some opportunities such as control of weeds, pests and diseases, and environmental benefits,
16 there are some challenges, including yield reduction, difficulty in pasture-cropping, grazing, and
17 groundcover maintenance in high rainfall zones, and development of persistent weeds and pests.
18 Adaptation measures based on agroecology entail enhancement of agrobiodiversity; improvement of
19 ecological processes and delivery of ecosystem services. They also entail strengthening of local
20 communities and recognition of the role and value of indigenous and local knowledge. Such practices
21 can enhance the sustainability and resilience of agricultural systems by buffering climate extremes,
22 reducing degradation of soils, and reversing unsustainable use of resources; outbreak of pests and
23 diseases and consequently increase yield without damaging biodiversity. Increasing and conserving
24 biological diversity such as soil microorganisms can promote high crop yields and sustain the
25 environment (Schmitz et al 2015; Bhattacharyya et al 2016; Garibaldi et al 2017).

26 Diversification of many components of the food system is a key element for increasing performance
27 and efficiency that may translate into increased resilience and reduced risks (integrated land
28 management systems, agrobiodiversity, indigenous and local knowledge, local food systems, dietary
29 diversity, the sustainable use of indigenous fruits, neglected and underutilised crops as a food source)
30 (*medium confidence*) (Makate et al. 2016; Lin 2011; Awodoyin et al. 2015).

31 The more diverse the food systems are, the more resilient they are in enhancing food security in the
32 face of biotic and abiotic stresses. Diverse production systems are important for providing regulatory
33 ecosystem services such as nutrient cycling, carbon sequestration, soil erosion control, reduction of
34 GHG emissions and control of hydrological processes (Chivenge et al. 2015). Further options for
35 adapting to change in both mean climate and extreme events are livelihood diversification (Michael
36 2017; Ford et al. 2015), and production diversity (Sibhatu et al. 2015).

37 Crop diversification, maintaining local genetic diversity, animal integration, soil organic matter
38 management, water conservation, and harvesting the role of microbial assemblages. These types of
39 farm management significantly affect communities in soil, plant structure, and crop growth in terms
40 of number, type, and abundance of species (Morrison-Whittle et al. 2017). Complementary strategies
41 towards sustainable agriculture (ecological intensification, strengthening existing diverse farming
42 systems and investment in ecological infrastructure) also address important drivers of pollinator
43 decline (IPBES 2016).

44 Evidence also shows that, together with other factors, on-farm agricultural diversity can translate into
45 dietary diversity at the farm level and beyond (Pimbert and Lemke 2018; Kumar et al. 2015; Sibhatu
46 et al. 2015). Dietary diversity is important but not enough as an adaptation option, but results in

1 positive health outcomes by increasing the variety of healthy products in people's diets and reducing
2 exposure to unhealthy environments.

3 Locally developed seeds and the concept of seed sovereignty can both help protect local
4 agrobiodiversity and can often be more climate resilient than generic commercial varieties (Watt
5 2016; Coomes et al., 2015; van Niekerk and Wynberg 2017; Vasconcelos et al. 2013). Seed exchange
6 networks and banks protect local agrobiodiversity and landraces, and can provide crucial lifelines
7 when crop harvests fail (Coomes et al. 2015; van Niekerk and Wynberg 2017; Vasconcelos et al.
8 2013).

9 Related to locally developed seeds, neglected and underutilised species (NUS) can play a key role in
10 increasing dietary diversity (*high confidence*) (Baldermann et al. 2016; van der Merwe et al. 2016;
11 Kahane et al. 2013; Muhanji et al. 2011) (see Box 5.3). These species can also improve nutritional and
12 economic security of excluded social groups, such as tribals (Nandal and Bhardwaj 2014; Ghosh-
13 Jerath et al. 2015), indigent (Kucich and Wicht 2016) or rural populations (Ngadze et al. 2017).

14

15 **Box 5.3 Climate change and indigenous food systems in the Hindu-Kush Himalayan** 16 **Region**

17 Diversification of production systems through promotion of Neglected and Underutilised Species
18 (NUS; also known as understudied, neglected, orphan, lost or disadvantaged crops) offers adaptation
19 opportunities to climate change, particularly in mountains. Neglected and Underutilised Species
20 (NUS) have a potential to improve food security and at the same time help protect and conserve
21 traditional knowledge and biodiversity. Scaling-up NUS requires training farmers and other
22 stakeholders on ways to adopt adequate crop management, quality seed, select varieties, farming
23 systems, soil management, development of new products, and market opportunities (Padulosi et al.
24 2013). Farmers in the Rasuwa district, in the mid-hills of Nepal, prefer to cultivate local bean, barley,
25 millet and local maize, rather than commodity crops because they are more tolerant to water stress and
26 extremely cold conditions (Adhikari et al. 2017). Farmers in the high-altitude cold climate of Nepal
27 prefer local barley with its short growing period because of a shorter growing window. Buckwheat is
28 commonly grown in the Hindu-Kush Himalayan (HKH) region mainly because it grows fast and
29 suppresses weeds. In Pakistan, quinoa (*Chenopodium quinoa*) grew and produced well under saline
30 and marginal soil where other crops would not grow (Adhikari et al. 2017).

31 At the same time, in many parts of the HKH region, a substantial proportion of the population is
32 facing malnutrition. Various factors are responsible for this, and lack of diversity in food and nutrition
33 resulting from production and consumption of few crops is one of them. In the past, food baskets in
34 this region consisted of many different edible plant species, many of which are now neglected and
35 underutilised. This is because almost all the efforts of the Green Revolution after 1960 focused on
36 major crops. Four crops viz. rice, wheat, maize and potato account for about 60% of global plant-
37 derived energy supply (Padulosi et al. 2013).

38 While the Green Revolution technologies substantially increased the yield of few crops and allowed
39 countries to reduce hunger, they also resulted in inappropriate and excessive use of agrochemicals,
40 inefficient water use, loss of beneficial biodiversity, water and soil pollution and significantly reduced
41 crop and varietal diversity. With farming systems moving away from subsistence-based to
42 commercial farming, farmers are also reluctant to grow these local crops because of low return, poor
43 market value and lack of knowledge about their nutritional environmental value.

44 However, transition from traditional diets based on local foods to a commercial crop-based diet with
45 high fats, salt, sugar and processed foods, increased the incidence of non-communicable diseases,
46 such as diabetes, obesity, heart diseases and certain types of cancer (Abarca-Gómez et al. 2017; NCD-

1 RisC 2016b, 2017b). This ‘hidden hunger’ – enough calories, but insufficient vitamins - is
2 increasingly evident in mountainous communities including the HKH region.

3 Internationally, there is rising interest nowadays on NUS, not only because they present tremendous
4 opportunities for fighting poverty, hunger and malnutrition, but also because of their role in mitigating
5 climate risk in agricultural production systems. NUS play an important role in mountain agro-
6 ecosystems because mountain agriculture is generally low-input agriculture, for which many NUS are
7 well adapted.

8 In the HKH region, mountains are agro-ecologically suitable for cultivation of traditional food crops,
9 such as barley, millet, sorghum, buckwheat, bean, grams, taro, yam and a vast range of wild fruits,
10 vegetables and medicinal plants. In one study carried out in two villages of mid-hills in Nepal, Khanal
11 et al. (2015) reported 52 indigenous crop species belonging to 27 families with their various uses.
12 Farming communities continue to grow various indigenous crops, albeit in marginal land, because of
13 their value on traditional food and associated culture. Nepal Agricultural Research Council (NARC)
14 has identified a list of indigenous crops based on their nutritional, medicinal, cultural and other
15 values.

16 Many indigenous crops supply essential micronutrients to the human body, and need to be conserved
17 in mountain food systems. Farmers in HKH region are cultivating and maintaining various indigenous
18 crops such as amaranthus, barley, black gram, horse gram, olarum, yam, rayo, sesame, niger, etc.
19 because of their nutritional value. Most of these indigenous crops are comparable with commercial
20 cereals in terms of dietary energy and protein content, but are also rich in micronutrients. For
21 example, pearl millet has higher content of calcium, iron, zinc, fiboflavin and folic acid than rice or
22 maize (Adhikari et al. 2017).

23 NUS can provide both climate resilience and more options for dietary diversity to the farming
24 communities of mountain ecosystems. Some of these indigenous crops have high medical importance.
25 For example, mountain people in the HKH region have been using *jammun* (i.e., *Syzygium cumini*) to
26 treat diabetes. In the Gilgit-Baltistan province of Pakistan, realising the importance of sea-buckthorn
27 for nutritional and medicinal purposes, local communities have expanded its cultivation to larger
28 areas. Many of these crops can be cultivated in marginal and/or fallow land which otherwise remains
29 fallow. Most of these species are drought resistant and can be easily grown in rainfed conditions in
30 non-irrigated land.

31
32 Dietary diversity has also been correlated (*medium evidence, medium agreement*) to agricultural
33 diversity in small-holder and subsistence farms (Ayenew et al. 2018; Jones et al. 2014; Jones 2017;
34 Pimbert and Lemke 2018), including both crops and animals, and has been proposed as a strategy to
35 reduce micronutrient malnutrition in developing countries (Tontisirin et al. 2002). In this regard, the
36 capacity of subsistence farming to supply essential nutrients in reasonable balance to the people
37 dependent on them has been considered as a means of overcoming their nutrient limitations in sound
38 agronomic and sustainable ways (Graham et al. 2007).

39 *Ecosystem-based adaptation (EbA)*. EbA is a set of nature-based methods addressing climate change
40 adaptation and food security by strengthening and conserving natural functions, goods and services
41 that benefit to people. EbA approaches to address food security provide co-benefits such as
42 contributions to health and improved diet, sustainable land management, economic revenue and water
43 security. EbA practices can reduce greenhouse gas emissions and increase carbon storage (USAID
44 2017).

45 For example, agroforestry systems can contribute to improving food productivity while enhancing
46 biodiversity conservation, ecological balance and restoration under changing climate conditions

1 (Mbow et al. 2014a; Paudela et al. 2017; Newaj et al. 2016; Altieri et al. 2015). Agroforestry systems
2 have been shown to reduce erosion through their canopy cover and their contribution to the micro-
3 climate and erosion control (Sida et al. 2018). Adoption of conservation farming practices such as
4 removing weeds from and dredging irrigation canals, draining and levelling land, and using organic
5 fertilisation were among the popular conservation practices in small-scale paddy rice farming
6 community of northern Iran (Ashoori and Sadegh 2016).

7 Adaptation potential of ecologically-intensive systems includes also forests and rivers ecosystems,
8 where improved resources management such as soil conservation, water cycling and agro-biodiversity
9 support the function of food production affected by severe climate change (Muthee et al. 2017). The
10 use of non-crop plant resources in agro-ecosystems (permaculture, perennial polyculture) can improve
11 ecosystem conservation and may lead to increased crop productivity (Balzan et al. 2016; Crews et al.
12 2018; Toensmeier 2016).

13 In summary, increasing the resilience of the food system through agroecology and diversification is an
14 effective way to achieve climate change adaptation (*robust evidence, high agreement*). Diversification
15 in the food system is a key adaptation strategy to reduce risks (e.g., implementation of integrated
16 production systems at landscape scales, broad-based genetic resources, and heterogeneous diets)
17 (*medium confidence*).

19 **5.3.2.4 Role of cultural values**

20 Food production and consumption are strongly influenced by cultures and beliefs. Culture, values and
21 norms are primary factors in most climate change and food system policies. The benefits of
22 integrating cultural beliefs and indigenous and local knowledge (ILK) into formal climate change
23 mitigation and adaptation strategies can add value to the development of sustainable climate change
24 that are rich in local aspirations, and planned with and for local people (Nyong et al. 2007).

25 Cultural dimensions are important in understanding how societies establish food production systems
26 and respond to climate change, since they help to explain differences in responses across populations
27 to the same environmental risks (Adger et al. 2013). There is an inherent adaptability of indigenous
28 people who are particularly connected to land use, developed for many centuries to produce specific
29 solutions to particular climate change challenges. Acknowledging that indigenous cultures across the
30 world are supporting many string strategies and beliefs that offer sustainable systems with pragmatic
31 solutions will help move forward the food and climate sustainability policies. For instance, in the
32 Sahel, the local populations have developed and implemented various adaptation strategies that
33 sustain their resilience despite many threats (Nyong et al. 2007). There is an increased consideration
34 of these local knowledge and cultural values and norms in the design and implementation of modern
35 mitigation and adaptation strategies.

36 There are some entrenched cultural beliefs and values that may be barriers to climate change
37 adaptation. For instance, culture has been shown to be a major barrier to adaptation for the Fulbe
38 ethnic group of Burkina Faso (Nielsen and Reenberg 2010). Thus, it is important to understand how
39 beliefs, values, practices and habits interact with the behaviour of individuals and collectivities that
40 have to confront climate change (Heyd and Thomas 2008). Granderson (2014) suggests that making
41 sense of climate change and its responses at the community level demands attention to the cultural
42 and political processes that shape how risk is conceived, prioritised and managed. For a discussion of
43 gender issues related to climate change, see Section 5.2.

44 Culturally sensitive risk analysis can deliver a better understanding of what climate change means for
45 society (O'Brien and Wolf 2010; Persson et al. 2015) and thus, how to better adapt. Murphy et al.
46 (2016) stated that culture and beliefs play an important role in adaptive capacity but that they are not

1 static. In the work done by Elum et al. (2017) in South Africa about farmers perception of climate
2 change, they concluded that perceptions and beliefs often have negative effects on adaptation options.

3 Culture is a key issue in food systems and the relation of people with nature. Food is an intrinsically
4 cultural process: food production shapes landscapes, which are in turn linked to cultural heritages and
5 identities (Koohafkan and Altieri 2011; Fuller and Qingwen 2013), and food consumption has a
6 strong cultural dimension. The loss of subsistence practices in modern cultures and its related
7 indigenous and local knowledge, has resulted in a loss of valuable adaptive capacities (Hernández-
8 Morcillo et al. 2014). This is so because these systems are often characterised by livelihood strategies
9 linked to the management of natural resources that that have been evolved to reduce overall
10 vulnerability to climate shocks (‘adaptive strategies’) and to manage their impacts ex-post (‘coping
11 strategies’) (Morton 2007; López-i-Gelats et al. 2016).

13 **5.3.3 Supply-side adaptation**

14 Supply-side adaptation takes place in the production (of crops, livestock, and aquaculture), storage,
15 transport, processing, and trade of food.

17 **5.3.3.1 Crop production**

18 There are many current agricultural management practices that can be optimised and scaled up to
19 advance adaptation. Among the often-studied adaptation options include increased soil organic matter,
20 improved cropland management, increased food productivity, prevention and reversal of soil erosion
21 (see Chapter 6 for evaluation of these practices in regard to desertification and land degradation).
22 Many analyses have demonstrated the effectiveness of soil management and changing sowing date,
23 crop type or variety (Waongo et al. 2015; Bodin et al. 2016; Teixeira et al. 2017; Waha et al. 2013;
24 Zimmermann et al. 2017; Chalise and Naranpanawa 2016; Moniruzzaman 2015; Sanz et al. 2017).
25 Biophysical adaptation options also include pest and disease management (Lamichhane et al. 2015)
26 and water management (Palmer et al. 2015; Korbel'ová and Kohnová 2017).

27 In Africa, Scheba (2017) found that conservation agriculture techniques were embedded in an
28 agriculture setting based on local traditional knowledge, including crop rotation, no or minimum
29 tillage, mulching, and cover crops. Cover cropping and no-tillage also improved soil health in a highly
30 commercialised arid irrigated system in California’s San Joaquin Valley, US (Mitchell et al. 2017).
31 Biofertilisers can enhance rice yields (Kantachote et al. 2016), and Amanullah and Khalid (2016)
32 found that manure and biofertiliser improve maize productivity under semi-arid conditions.

33 Adaptation also involves use of current genetic resources as well as breeding programs for both crops
34 and livestock. More drought, flood and heat-resistant crop varieties (Atlin et al. 2017; Mickelbart et
35 al. 2015; Singh et al. 2017) and improved nutrient and water use efficiency, including overabundance
36 as well as water quality (such as salinity) (Bond et al. 2018) are aspects to factor in to the design of
37 adaptation measures. Both availability and adoption of these varieties is a possible path of adaptation
38 and can be facilitated by new outreach policy and capacity building.

39 Water management is another key area for adaptation. Increasing water availability and reliability of
40 water for agricultural production using different techniques of water harvesting, storage, and its
41 judicious utilisation through farm ponds, dams, and community tanks in rainfed agriculture areas have
42 been presented by Rao et al. (2017) and Rivera-Ferre et al. (2016a). In addition, improved drainage
43 systems (Thiel et al. 2015), and Alternate Wetting and Drying (AWD) techniques for rice cultivation
44 (Howell et al. 2015; Rahman and Bulbul 2015) have been proposed. Efficient irrigation systems have
45 been also analysed and proposed by (Jägermeyr et al. 2016) Naresh et al (2017) (Gunarathna et al.

1 2017; Chartzoulakis and Bertaki 2015). Recent innovation includes using farming system with low
2 usage of water such as drip-irrigation or hydroponic systems mostly in urban farming.

3 4 **5.3.3.2 Livestock production systems**

5 Considering the benefits of higher temperature in temperate climate and the increase of pasture with
6 incremental warming in some humid and temperate grasslands, as well as potential negative effects,
7 can be useful in planning adaptation strategies to future climate change. Rivera-Ferre et al. (2016b)
8 characterize adaptation for different livestock systems as managerial, technical, behavioural and
9 policy-related options. Managerial included production adjustments (e.g., intensification, integration
10 with crops, shifting from grazing to browsing species, multispecies herds, mobility, soil and nutrient
11 management, water management, pasture management, corralling, feed and food storage, farm
12 diversification or cooling systems); and changes in labor allocation (diversifying livelihoods, shifting
13 to irrigated farming, labor flexibility). Technological options included breeding strategies and
14 information technology research. Behavioral options are linked to cultural patterns and included
15 encouraging social collaboration and reciprocity, e.g., livestock loans, communal planning, food
16 exchanges. and information sharing. Policy options are discussed in Section 5.7 and Chapter 7.

17 18 **5.3.3.3 Aquaculture, fisheries, and agriculture interactions**

19 Options may include livelihood diversification within and across sectors of fisheries, aquaculture and
20 agriculture. Thus, adaptation options need to provide management approaches and policies that build
21 the livelihood asset base, reducing vulnerability to multiple stressors with a multi-sector perspective
22 (Badjeck et al. 2010). In Bangladesh fishing pressure on post-larval prawns has increased as displaced
23 farmers have shifted to fishing following salt-water intrusion of agricultural land (Ahmed et al. 2013).
24 In West Africa, strategies to cope with sudden shifts in fisheries are wider-reaching and have included
25 turning to seafood import (Gephart et al. 2017) or terrestrial food production including farming and
26 bush-meat hunting on land (Brashares et al. 2004). Proposed actions for adaptation include effective
27 governance, improved management and conservation, efforts to maximise societal and environmental
28 benefits from trade, increased equitability of distribution and innovation in food production, and the
29 continued development of low-input and low-impact aquaculture (FAO 2018c).

30 Particular adaptation strategies proposed by FAO (2014a) include diverse and flexible livelihood
31 strategies, such as introduction of fish ponds in areas susceptible to intermittent flood/drought periods;
32 flood-friendly small-scale homestead bamboo pens with trap doors allowing seasonal floods to occur
33 without loss of stocked fish; cage fish aquaculture development using plankton feed in reservoirs
34 created by dam building; supporting the transition to different species, polyculture and integrated
35 systems, allowing for diversified and more resilient systems; promotion of rice–fish farming systems
36 reducing overall water needs and providing integrated pest management; and supporting transitions to
37 alternative livelihoods.

38 Risk reduction initiatives include innovative weather-based insurance schemes being tested for
39 applicability in aquaculture and fisheries and climate risk assessments introduced for integrated
40 coastal zone management. For aquaculture’s contribution to building resilient food systems, Troell et
41 al. (2014) found that aquaculture could potentially enhance resilience through improved resource use
42 efficiencies and increased diversification of farmed species, locales of production, and feeding
43 strategies. Yet, given its high reliance on terrestrial crops and wild fish for feeds, its dependence on
44 freshwater and land for culture sites and its environmental impacts reduce this potential. For instance,
45 the increase in aquaculture worldwide may enhance land competition for feed crops, increasing price
46 levels and volatility and worsening food insecurity among the most vulnerable populations.

1 **5.3.3.4 Transport and storage**

2 Fewer studies have been done on adaptation of food system transport and storage compared to the
3 many studies on adaptation to climate in food production.

4 *Transport.* One transport example is found in Bangkok. Between mid-November 2011 and early
5 January 2012, Bangkok, the capital city of Thailand, faced its most dramatic flood in approximately
6 70 years with most transport networks cut-off or destroyed. This caused large-scale disruption of the
7 national food supply chains since they were centrally organised in the capital city (Allen et al. 2017).
8 From this experience, the construction and management of ‘climate-proof’ rural roads and transport
9 networks is argued as one the most important adaptation strategies for climate change and food
10 security in Thailand (Rattanachot et al. 2015).

11 Similarly in Africa, it has been shown that enhanced transportation networks combined with other
12 measures could reduce the impact of climate change on food and nutrition security (Brown et al.
13 2017b). This suggests that strengthening infrastructure and logistics for transport would significantly
14 enhance resilience to climate change while improving food and nutrition security in developing
15 counties.

16 *Storage.* Storage refers to both structures and technologies for storing seed as well as produce.
17 Predominant storage methods used in Uganda are single-layer woven polypropylene bags (popularly
18 called “kavera” locally), chemical insecticides and granaries. Evidence from Omotilewa et al. (2018)
19 showed that the introduction of new storage technology called Purdue Improved Crop Storage (PICS)
20 could contribute to climate change adaptation. PICS is a chemical-free airtight triple-layered
21 technology consisting of two high-density polyethylene inner liners and one outer layer of woven
22 polypropylene bag. Its adoption has increased the number of households planting hybrid maize
23 varieties that are more susceptible to insect pests in storage than traditional lower-yielding varieties.
24 Such innovations could help to protect crops more safely and for longer periods from postharvest
25 insect pests that are projected to increase as result of climate change, thus contributing to food
26 security. In the Indo-Gangetic Plains many different storage structures based on ILK provide reliable
27 and low-cost options made of local materials. For example, elevated grain stores protect from floods,
28 but also provide for air circulation to prevent rot and to control insects and other vermin (Rivera-Ferre
29 et al. 2013).

30

31 **5.3.3.5 Trade and processing**

32 Adaptation measures are also being considered in trade, processing, and packaging, other components
33 of the food system. These will enable availability, stability, and safety of food under changing climate
34 conditions.

35 *Trade.* Brooks & Matthews (2015) found that food trade increases the availability of food by enabling
36 products to flow from surplus to deficit areas, raises incomes and favors access to food, improves
37 utilisation by increasing the diversity of national diets while pooling production risks across
38 individual markets to maintain stability.

39 *Processing.* Growth of spoilage bacteria of red meat and poultry during storage due to increasing
40 temperature has been demonstrated by European Food Safety Authority (EFSA Panel on Biological
41 Hazards 2016). In a recent experiment conducted on the optimisation of processing conditions of
42 Chinese traditional smoke-cured bacon (Larou), Liu et al. (2018a) showed that the use of new natural
43 coating solution composed of lysozyme, sodium alginate, and chitosan during storage period resulted
44 in 99.69% rate of reducing deterioration after 30-day storage. Also, the use of High Hydrostatic
45 Pressure (HHP) technology to inactivate pathogenic, spoilage microorganisms and enzymes with little
46 or no effects on the nutritional and sensory quality of foods have been described by Wang et al.

1 (2016) and Ali et al. (2018) as new advances in processing and packaging fruits, vegetables, meats,
2 seafood, dairy, and egg products.

3 In summary, there are many practices that can be optimised and scaled up to advance supply-side
4 adaptation. On-farm adaptation options include increased soil organic matter and erosion control in
5 cropland, improved livestock and grazing land management, and transition to different species,
6 polyculture and integrated systems in aquaculture. Crop and livestock genetic improvements include
7 tolerance to heat, drought, and pests and diseases. Food transport, storage, trade, and processing will
8 likely play increasingly important roles in adapting to climate change-induced food insecurity.

9

10 **5.3.4 Demand-side adaptation**

11 Adaptation in the demand side of the food system involves consumption practices, diets, and reducing
12 food loss and waste. Recent studies showed that supply-side adaptation measures alone will not be
13 sufficient to sustainably achieve food security under climate change (Springmann et al. 2018b;
14 Swinburn et al. 2019; Bajželj et al. 2014). As noted by Godfray (2015), people with higher income
15 demand more varied diets, and typically ones that are richer in meat and other food types that require
16 more resources to produce. Therefore, both supply-side (production, processing, transport, trade, etc.)
17 and demand-side solutions (changing diets, food loss and waste reduction, etc.) can be effective in
18 adapting to climate change (Creutzig et al. 2016) (see Section **Error! Reference source not found.**
19 or food loss and waste).

20 The implications of dietary choice can have severe consequences for land. For example, Alexander et
21 al. 2016, found that if every country were to adopt the UK's 2011 average diet and meat consumption,
22 95% of global habitable land area would be needed for agriculture – up from 50% of land currently
23 used. For the average US diet, 178% of global land would be needed (relative to 2011) (Alexander et
24 al. 2016); and for “business as usual” dietary trends and existing rates of improvement in yields, 55%
25 more land would be needed above baseline (2009) (Bajželj et al. 2014). Changing dietary habits has
26 been suggested as an effective food route to affect land use (Beheshti et al. 2017) and promote
27 adaptation to climate change through food demand.

28 Most literature has focused on demand-side options that analyse the effects on climate change
29 mitigation by dietary changes. Little focus has been brought on demand-side adaptation measures to
30 adjust the demand to the food challenges related to drivers such as market, climate change, inputs
31 limitations (e.g., fossil fuels, nitrogen, phosphorus), food access, and quality. Adding to that, the high
32 cost of nutritious foods contributes to a higher risk of overweight and obesity (FAO 2018d).
33 Adaptation measures relate also to the implications of easy access to inexpensive, high-calorie, low-
34 nutrition foods which have been shown to lead to malnutrition (Section 5.1). Therefore, adaptation
35 related to diet may be weighed against the negative side-effects on health of current food choices.

36 Reduction in the demand for animal-based food products and increasing proportions of plant-based
37 foods in diets, particularly pulses and nuts; and replacing red meat with other more-efficient protein
38 sources are demand-side adaptation measures (Machovina et al. 2015) (see also Section 5.5.2). For
39 example, replacing beef in the US diet with poultry can meet caloric and protein demands of about
40 120 to 140 million additional people consuming the average American diet (Shepon et al. 2016).
41 Similar suggestions are made for adopting the benefits of moving to plant-based protein, such as
42 beans (Harwatt et al. 2017).

43 The main reason why reducing meat consumption is an adaptation measure is because it reduces
44 pressure on land and water and thus our vulnerability to climate change and inputs limitations
45 (Vanham et al. 2013). For animal feed, ruminants can have positive ecological effects (species
46 diversity, soil carbon) if they are fed extensively on existing grasslands. Similarly, reducing waste at

1 all points along the entire food chain is a significant opportunity for improving demand-side
2 adaptation measures (Godfray 2015).

3 It is important to highlight the opportunities for improving the feed-to-meat conversion considered as
4 a form of food loss. However, the unique capacity of ruminants to produce high-quality food from
5 low-quality forage, in particular from landscapes that cannot be cropped and from cellulosic biomass
6 that humans cannot digest could be seen as an effective way to improve the feed:meat ratio (Cawthorn
7 and Hoffman 2015).

8 In summary, there is potential for demand-side adaptation, such as adoption of diets low in animal-
9 sourced products, in conjunction with reduction in food loss and waste to contribute to reduction in
10 food demand, land sparing, and thus need for adaptation.

11

12 **5.3.5 Institutional measures**

13 To facilitate the scaling up of adaptation throughout the food system, institutional measures are
14 needed at global, regional, national, and local levels (See Section 5.7). Institutional aspects including
15 policies and laws depend on scale and context. International institutions (financial and policies) are
16 driving many aspects of global food systems (e.g., UN agencies, international private sector
17 agribusinesses and retailers). Many others operate at local level and strongly influence livelihoods and
18 markets of smallholder farmers. Hence, differentiation in the roles of the organisations, their missions
19 and outcomes related to food and climate change action need to be clearly mapped and understood.

20 Awareness about the institutional context within which adaptation planning decisions are made is
21 essential for the usability of climate change projection (Lorenz 2017) (Chapter 7 SRCCL). In the
22 planning and operational process of food production, handling and consumption, the environment
23 benefits and climate change goals can be mainstreamed under sustainable management approaches
24 that favor alternative solutions for inputs, energy consumption, transformation and diet. For instance,
25 land use planning would guide current and future decision making and planners in exploring
26 uncertainty to increase the resilience of communities (Berke & Stevens 2016). One of the important
27 policy implications for enhanced food security are the trade-offs between agricultural production and
28 environmental concerns, including the asserted need for global land use expansion, biodiversity and
29 ecological restoration (See Section 5.6) (Meyfroidt 2017).

30 There are a number of adaptation options in agriculture in the form of policy, planning, governance
31 and institutions (Lorenz 2017). For example, early spatial planning action is crucial to guide decision-
32 making processes and foster resilience in highly uncertain future climate change (Brunner and Grêt-
33 Regamey, 2016). Institutions may develop new capacities to empower value chain actors take climate
34 change into account as they develop quality products, promote adoption of improved diet for healthier
35 lifestyles, aid the improvement of livelihoods of communities, and further socioeconomic
36 development (Sehmi et al. 2016). Other adaptation policies include property rights and land tenure
37 security as legal and institutional reforms to ensure transparency and access to land that could
38 stimulate adaptation to climate change Antwi-Agyei et al. (2015).

39

40 **5.3.5.1 Global initiatives**

41 Climate change poses serious wide-ranging risks, requiring a broader approach in fighting the
42 phenomenon. The United Nations Framework Convention on Climate Change (UNFCCC) and its
43 annual Conferences of the Parties (COPs) has been instrumental in ensuring international cooperation
44 in the field of tackling the impacts of climate change in a broader framework (Cléménçon 2016). The
45 National Adaptation Plan (NAP) program under the UNFCCC, was established to: identify vulnerable
46 regions; assess the impacts of climate change on food security; and prioritise adaptation measures for

1 implementation to increase resilience. The National Adaptation Programs of Action (NAPAs) was
2 also established to support least-developed countries (LDCs) address their particular challenges in
3 adaptation, to enhance food security among other priorities. The Paris Agreement (UNFCCC 2015) is
4 a major victory for small island states and vulnerable nations that face climate change-related impacts
5 of floods and droughts resulting in food security challenges. Adaptation and mitigation targets set by
6 the parties through their nationally determined commitments (NDCs) are reviewed internationally to
7 ensure consistency and progress towards actions (Falkner 2016).

8 The Food and Agriculture Organization of the United Nations (FAO) also plays a significant role in
9 designing and coordinating national policies to increase adaptation and food security. The five key
10 strategic objectives of FAO (Help eliminate hunger, food insecurity and malnutrition; Make
11 agriculture, forestry and fisheries more productive and sustainable; Reduce rural poverty; Enable
12 inclusive and efficient agricultural and food systems; and increase the resilience of livelihoods to
13 climate threats) (FAO 2018e), all relate to building resilience and increasing global adaptation to
14 climate variability.

15 In support of the Paris Agreement, FAO launched a global policy, “Tracking Adaptation” with the
16 aim of monitoring the adaptation processes and outcomes of the parties to increase food security, and
17 make available technical information for evaluation by stakeholders. In response to the estimated
18 world population of 9.7 billion by 2050, FAO adopted the Climate Smart Agriculture (CSA) approach
19 to increase global food security without compromising environmental quality (See Section 5.6). FAO
20 supports governments at the national level to plan CSA programs and to seek climate finance to fund
21 their adaptation programs.

22 The Global Commission on Adaptation, co-managed by World Resources Institute (WRI) and the
23 Global Center on Adaptation, seeks to accelerate adaptation action by elevating the political visibility
24 of adaptation and focusing on concrete solutions (Global Commission on Adaptation 2019). The
25 Commission works to demonstrate that adaptation is a cornerstone of better development, and can
26 help improve lives, reduce poverty, protect the environment, and enhance resilience around the world.
27 The Commission is led by Ban Ki-moon, 8th Secretary-General of the United Nations, Bill Gates, co-
28 chair of the Bill & Melinda Gates Foundation, and Kristalina Georgieva, CEO, World Bank. It is
29 convened by 17 countries and guided by 28 commissioners. A global network of research partners and
30 advisors provide scientific, economic, and policy analysis.

32 **5.3.5.2 National policies**

33 The successful development of food systems under climate change conditions requires a national-
34 level management that involves the cooperation of a number of institutions and governance entities to
35 enable more sustainable and beneficial production and consumption practices.

36 For example, Nepal has developed a novel multi-level institutional partnership, under the Local
37 Adaptation Plan of Action (LAPA), which is an institutional innovation that aims to better integrate
38 local adaptation planning processes and institutions into national adaptation processes. That includes
39 collaboration with farmers and other non-governmental organisations (Chhetri et al. 2012). By
40 combining conventional technological innovation process with the tacit knowledge of farmers, this
41 new alliance has been instrumental in the innovation of location-specific technologies thereby
42 facilitating the adoption of technologies in a more efficient manner.

43 National Adaptation Planning of Indonesia was officially launched in 2014 and was an important
44 basis for ministries and local governments to mainstream climate change adaptation into their
45 respective sectoral and local development plans Kawanishi et al. (2016). Crop land use policy to
46 switch from crops that are highly impacted by climate change to those that are less vulnerable were
47 suggested for improving climate change adaptation policy processes and outcomes in Nepal (Chalise

1 and Naranpanawa 2016). Enhancement of representation, democratic and inclusive governance, as
2 well as equity and fairness for improving climate change adaptation policy processes and outcomes in
3 Nepal were also suggested as intuitional measure by Ojha et al. (2015). Further, food, nutrition, and
4 health policy adaptation options such as social safety nets and social protection have been
5 implemented in India, Pakistan, Middle East and North Africa (Devereux 2015; Mumtaz and
6 Whiteford 2017; Narayanan and Gerber 2017).

7 Financial incentives policies at the national scale used as adaptation options include taxes and
8 subsidies; index-based weather insurance schemes; and catastrophe bonds (Zilberman et al. 2018;
9 Linnerooth-Bayer and Hochrainer-Stigler 2015; Ruiter et al. 2017; Campillo et al. 2017).
10 Microfinance, disaster contingency funds, and cash transfers are other mechanisms (Ozaki 2016;
11 Kabir et al. 2016).

12

13 **5.3.5.3 Community-based adaptation**

14 Community-based adaptation (CBA) builds on social organisational capacities and resources to
15 addressing food security and climate change. CBA represents bottom-up approaches and localised
16 adaptation measures where social dynamics serve as the power to respond to the impacts of climate
17 change (Ayers and Forsyth 2009). It identifies, assists, and implements development activities that
18 strengthen the capacity of local people to adapt to living in a riskier and less predictable climate,
19 while ensuring their food security.

20 Klenk et al. (2017) found that mobilisation of local knowledge can inform adaptation decision-making
21 and may facilitate greater flexibility in government-funded research. As an example, rural innovation
22 in terrace agriculture developed on the basis of a local coping mechanism and adopted by peasant
23 farmers in Latin America may serve as an adaptation option to climate change (Bocco and
24 Napoletano, 2017). Clemens et al. (2015) indicated that learning alliances provided social learning
25 and knowledge-sharing in Vietnam through an open dialogue platform that provided incentives and
26 horizontal exchange of ideas.

27 Community-based adaptation generates strategies through participatory processes, involving local
28 stakeholders and development and disaster risk–reduction practitioners. Fostering collaboration and
29 community stewardship is central to the success of CBA (Scott et al. 2017). Preparedness behaviours
30 that are encouraged include social connectedness, education, training, and messaging; CBA also can
31 encompass beliefs that might improve household preparedness to climate disaster risk (Thomas et al.
32 2015). Reliance on social networks, social groups connectivities, or moral economies reflect the
33 importance of collaboration within communities (Reuter 2018; Schramski et al. 2017).

34 Yet, community-based adaptation also needs to consider methods that engage with the drivers of
35 vulnerability as part of community-based approaches, particularly questions of power, culture,
36 identity and practice (Ensor et al. 2018). The goal is to avoid maladaptation or exacerbation of
37 existing inequalities within the communities (Buggy and McNamara 2016). For example, in the
38 Pacific Islands, elements considered in a CBA plan included people’s development aspirations;
39 immediate economic, social and environmental benefits; dynamics of village governance, social rules
40 and protocols; and traditional forms of knowledge that could inform sustainable solutions (Remling
41 and Veitayaki 2016).

42 With these considerations, community-based adaptation can help to link local adaptation with
43 international development and climate change policies (Forsyth 2013). In developing CBA programs,
44 barriers exist that may hinder implementation. These include poor coordination within and between
45 organisations implementing adaptation options, poor skills, poor knowledge about climate change,
46 and inadequate communication among stakeholders (Spires et al. 2014). A rights-based approach has

1 been suggested to address issues of equality, transparency, accountability and empowerment in
2 adaptation to climate change (Ensor et al. 2015).

3 In summary, institutional measures, including risk management, policies, and planning at global,
4 national, and local scales can support adaptation. Advance planning and focus on institutions can aid
5 in guiding decision-making processes and foster resilience. There is evidence that institutional
6 measures can support the scaling up of adaptation and thus there is reason to believe that systemic
7 resilience is achievable.

8

9 **5.3.6 Tools and finance**

10 **5.3.6.1 Early Warning Systems**

11 Many countries and regions in the world have adopted early warning systems (EWS) to cope with
12 climate variability and change as it helps to reduce interruptions and improve response times before
13 and after extreme weather events (Ibrahim and Kruczkiewicz 2016). The Early Warning and Early
14 Action (EW/EA) framework has been implemented in West Africa (Red Cross 2011) and
15 Mozambique (DKNC 2012). Bangladesh has constructed cyclone shelters where cyclone warnings are
16 disseminated and responses organised (Mallick et al. 2013). In Benin, a Standard Operating Procedure
17 is used to issue early warnings through the UNDP Climate Information and Early Warning Systems
18 Project (UNDP 2016).

19 However, there are some barriers to building effective early warning systems in Africa, such as lack
20 of reliable data and distribution systems, lack of credibility, and limited relationships with media and
21 government agencies (UNDP 2016). Mainstreaming early warning systems in adaptation planning
22 could present a significant opportunity for climate disaster risk reduction (Zia and Wagner 2015).
23 Enenkel et al. (2015) suggested that the use of smartphone applications that concentrate on food and
24 nutrition security could help with more frequent and effective monitoring of food prices, availability
25 of fertilisers and drought-resistant seeds, and could help to turn data streams into useful information
26 for decision support and resilience building.

27 GIS and remote sensing technology are used for monitoring and risk quantification for broad-
28 spectrum stresses such as drought, heat, cold, salinity, flooding, and pests (Skakun et al. 2017; Senay
29 et al. 2015; Hossain et al. 2015; Brown 2016), while site-specific applications, such as drones, for
30 nutrient management, precision fertilisers, and residue management can help devise context-specific
31 adaptations (Campbell et al. 2016; Baker et al. 2016). Systematic monitoring and remote sensing
32 options, as argued by Aghakouchak et al. (2015), showed that satellite observations provide
33 opportunities to improve early drought warning. Waldner et al. (2015) found that cropland mapping
34 allows strategic food and nutrition security monitoring and climate modelling.

35 Access to a wide range of adaptation technologies for precipitation change is important, such as
36 rainwater harvesting, wastewater treatment, stormwater management and bioswales, water demand
37 reduction, water-use efficiency, water recycling and reuse, aquifer recharge, inter-basin water transfer,
38 desalination, and surface-water storage (ADB 2014).

39

40 **5.3.6.2 Financial resources**

41 Financial instruments such as micro-insurance, index-based insurance, provision of post-disaster
42 finances for recovery and pre-disaster payment are fundamental means to reduce lower and medium
43 level risks (Linnerooth-Bayer and Hochrainer-Stigler 2014). Fenton & Paavola, 2015; Dowla, 2018).
44 Hammill et al. (2010) found that microfinance services (MFS) are especially helpful for the poor.
45 MFS can provide poor people with the means to diversify, accumulate and manage the assets needed
46 to become less susceptible to shocks and stresses. As a result, MFS plays an important role in

1 vulnerability reduction and climate change adaptation among some of the poor. The provision of
2 small-scale financial products to low-income and otherwise disadvantaged groups by financial
3 institutions can serve as adaptation to climate change. Access to finance in the context of climate
4 change adaptation that focuses on poor households and women in particular is bringing encouraging
5 results (Agrawala and Carraro 2010).

6 In summary, effective adaptation strategies can reduce the negative impacts of climate change. Food
7 security under changing climate conditions depends on adaptation throughout the entire food system –
8 production, supply chain, and consumption/demand, as well as reduction of food loss and waste.
9 Adaptation can be autonomous, incremental, or transformative, and can reduce vulnerability and
10 enhance resilience. Local food systems are embedded in culture, beliefs and values, and indigenous
11 and local knowledge can contribute to enhancing food system resilience to climate change (*high*
12 *confidence*). Institutional and capacity-building measures are needed to scale up adaptation measures
13 across local, national, regional, and global scales.

15 **5.4 Impacts of food systems on climate change**

16 **5.4.1 Greenhouse gas emissions from food systems**

17 This chapter assesses the contributions of the entire food system to greenhouse gas (GHG) emissions.
18 Food systems emissions include CO₂ and non-CO₂ gases, specifically those generated from: i) crop
19 and livestock activities within the farm gate (Table 5.4, category ‘Agriculture’); ii) land use and land
20 use change dynamics associated with agriculture (Table 5.4, category ‘Land Use’); and iii) food
21 processing, retail and consumption patterns, including upstream and downstream processes such as
22 manufacture of chemical fertilisers and fuel (Table 5.4, category ‘Beyond Farm Gate’). The first two
23 categories comprise emissions reported by countries in the AFOLU (Agriculture, Forestry, and Other
24 Land Use) sectors of national GHG inventories; the latter comprises emissions reported in other
25 sectors of the inventory, as appropriate, for instance, industrial processes, energy use, and food loss
26 and waste.

27 The first two components (agriculture and land use) identified above are well quantified and
28 supported by an ample body of literature (Smith et al. 2014). During the period 2007-2016, global
29 agricultural non-CO₂ emissions from crop and livestock activities within the farm gate were 6.2 ± 1.9
30 Gt CO₂-eq yr⁻¹ during 2007-2016, with methane (142 ± 43 Mt CH₄ yr⁻¹, or 4.1 ± 1.2 Gt CO₂-eq yr⁻¹)
31 contributing in CO₂eq about twice as much as nitrous oxide (8.3 ± 2.3 Mt N₂O yr⁻¹, or 2.1 ± 0.6 Gt
32 CO₂-eq yr⁻¹) to this total (see Table 2.2 in Chapter 2). Emissions from land use associated with
33 agriculture in some regions, such as from deforestation and peatland degradation (both processes
34 involved in preparing land for agricultural use), added globally during the same period another $4.8 \pm$
35 2.4 Gt CO₂-eq yr⁻¹ (see Chapter 2). These estimates are associated with uncertainties of about 30%
36 (agriculture) and 50% (land use), as per IPCC AR5 (Smith et al. 2014).

37 Agriculture activities within the farm gate and associated land use dynamics are therefore responsible
38 for about 11.0 ± 3.1 Gt CO₂-eq yr⁻¹, or some 20% of total anthropogenic emissions (Table 5.4),
39 consistent with post-AR5 findings (e.g., Tubiello et al. (2015)). In terms of individual gases, the
40 contributions of agriculture to total emissions by gas are significantly larger. For instance, over the
41 period 2010-2016, methane gas emissions within the farm gate represented about half of the total CH₄
42 emitted by all sectors, while nitrous dioxide gas emissions within the farm gate represented about
43 three-quarters of the total N₂O emitted by all sectors (Tubiello 2019). In terms of carbon, CO₂

1 emissions from deforestation and peatland degradation linked to agriculture contributed about 10% of
2 the CO₂ emitted by all sectors in 2017 (Le Quéré et al. 2018).

3 Food systems emissions beyond the farm gate, such as those upstream from manufacturing of
4 fertilisers, or downstream such as food processing, transport and retail, and food consumption,
5 generally add to emissions from agriculture and land use, but their estimation is very uncertain due to
6 lack of sufficient studies. The IPCC AR5 (Fischedick et al. 2014) provided some information on these
7 other food system components, noting that emissions beyond the farm gate in developed countries
8 may equal those within the farm gate, and cited one study estimating world total food system
9 emissions to be up to 30% of total anthropogenic emissions (Garnett 2011). More recently, Poore and
10 Nemecek (2018), by looking at a database of farms and using a combination of modelling approaches
11 across relevant processes, estimated a total contribution of food systems around 26% of total
12 anthropogenic emissions. Total emissions from food systems may thus account for 25-30% of total
13 GHG emissions (*medium confidence*).

14 Based on the available literature, a break-down of individual contributions of food systems emissions
15 is show in Table 5.4, between those from agriculture within the farm gate (10-12%) (*high confidence*);
16 emissions from land use and land use change dynamics such as deforestation and peatland
17 degradation, which are associated with agriculture in many regions (8-10%) (*high confidence*); and
18 those from food supply chain activities past the farm gate, such as storage, processing, transport, and
19 retail (5-10%) (*limited evidence, medium agreement*). Note that the corresponding lower range of
20 emissions past the farm gate, i.e., 2.5 Gt CO₂-eq yr⁻¹ (Table 5.4), is consistent with recent estimates
21 made by Poore and Nemecek (2018). Contributions from food loss and waste are implicitly included
22 in these estimates of total emissions from food systems (See Section 5.5.2.5). They may account for
23 8–10% of total GHG emissions from agriculture and land use (FAO 2013b) (*low confidence*).

24

25 **Table 5.4 GHG emissions (Gt CO₂eq yr⁻¹) from the food system and their contribution (%) to total**
26 **anthropogenic emissions. Mean of 2007-2016 period.**

| Food system component | Emissions (Gt CO ₂ eq yr ⁻¹) | Share in mean total emissions (%) |
|-----------------------|--|---|
| Agriculture | 6.2 ± 1.9 ^a | 10-12% |
| Land use | 4.8 ± 2.4 ^a | 8-10% |
| Beyond farm gate | 3.8 ± 1.3 ^b | 5-10% |
| Food system (Total) | 14.8 ± 3.4 | 25-30% |

27 Notes: Food system emissions are estimated by combining emissions data from a) FAOSTAT (2018) and US
28 EPA (See also Chapter 2) and b) Garnett (2011) and Poore and Nemecek (2018). Percentage shares were
29 computed by using a total emissions value for the period 2007-2016 of nearly 51 Gt CO₂-eq yr⁻¹ (See Chapter 2).
30 GWP values used are those , and by using GWP values of the IPCC AR5 with no climate feedback (GWP-
31 CH4=28; GWP-N2O=265)..

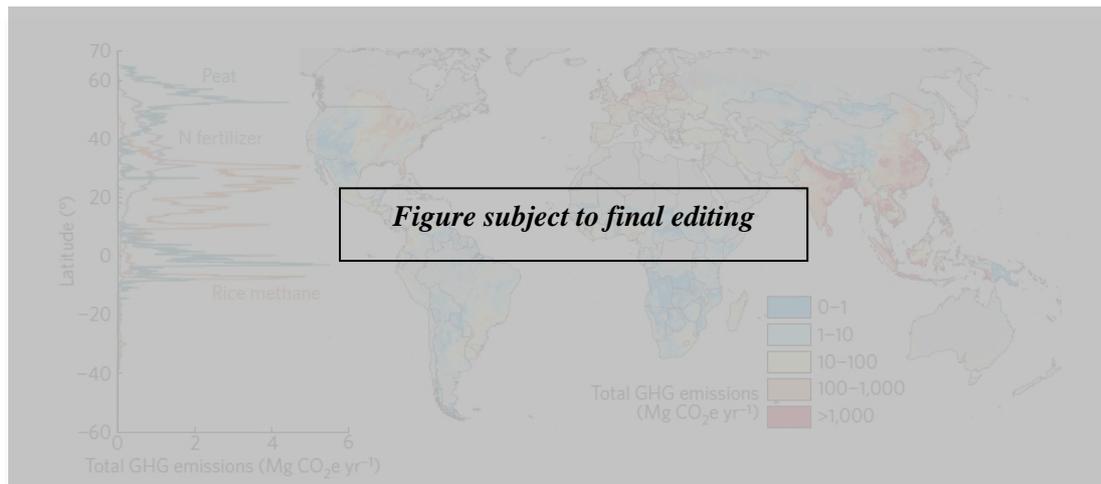
32

33 **5.4.2 Greenhouse gas emissions from croplands and soils**

34 Since AR5, a few studies have quantified separate contributions of crops and soils on the one hand,
35 and livestock on the other, to the total emissions from agriculture and associated land use. For
36 instance, Carlson et al. (2017) estimated emissions from cropland to be in the range of 2–3 GtCO₂-eq

1 yr⁻¹, including methane emissions from rice, CO₂ emissions from peatland cultivation, and N₂O
 2 emissions from fertiliser applications. Data from FAOSTAT (2018), recomputed to use AR5 GWP
 3 values, indicated that cropland emissions from these categories were 3.6 ± 1.2 Gt CO₂-eq yr⁻¹ over the
 4 period 2010–2016; two-thirds of this were related to peatland degradation, followed by N₂O
 5 emissions from synthetic fertilisers and methane emissions from paddy rice fields (Tubiello 2019).
 6 These figures are a subset of the total emissions from agriculture and land use reported in Table 5.4.
 7 Asia, especially India, China and Indonesia accounted for roughly 50% of global emissions from
 8 croplands. Figure 5.8 shows the spatial distribution of emissions from cropland according to Carlson
 9 et al. (2017), not including emissions related to deforestation or changes in soil carbon.

10



11

12 **Figure 5.8 Cropland GHGs consist of CH₄ from rice cultivation, CO₂, N₂O, and CH₄ from peatland**
 13 **draining, and N₂O from N fertiliser application. Total emissions from each grid cell are concentrated in**
 14 **Asia, and are distinct from patterns of production intensity (Carlson et al. 2017).**

15 5.4.3 Greenhouse gas emissions from livestock

16 Emissions from livestock include non-CO₂ gases from enteric fermentation from ruminant animals
 17 and from anaerobic fermentation in manure management processes, as well as non-CO₂ gases from
 18 manure deposited on pastures (Smith et al. 2014). Estimates after the AR5 include those from Herrero
 19 et al. (2016), who quantified non-CO₂ emissions from livestock to be in the range of 2.0–3.6 GtCO₂-
 20 eq yr⁻¹, with enteric fermentation from ruminants being the main contributor. FAOSTAT (2018)
 21 estimates of these emissions, renormalized to AR5 GWP values, were 4.1 ± 1.2 Gt CO₂-eq yr⁻¹ over
 22 the period 2010–2016.

23 These estimates of livestock emissions are for those generated within the farm gate. Adding emissions
 24 from relevant land use change, energy use, and transportation processes, FAO (2014a) and Gerber et
 25 al. (2013) estimated livestock emissions of up to 5.3 ± 1.6 GtCO₂-eq yr⁻¹ circa the year 2010 (data
 26 from original papers, but scaled to SAR global warming potential (GWP) values for methane, for
 27 comparability with previous results).

28 All estimates agree that cattle are the main source of global livestock emissions (65–77%). Livestock
 29 in low and middle-income countries contribute 70% of the emissions from ruminants and 53% from
 30 monogastric livestock (animals without ruminant digestion processes such as sheep, goats, pigs, and
 31 poultry), and these are expected to increase as demand for livestock products increases in these
 32 countries (Figure 5.9). In contrast to the increasing trend in absolute GHG emissions, GHG emissions
 33 intensities, defined as GHG emissions per unit produced, have declined globally and are about 60%
 34 lower today than in the 1960s. This is largely due to improved meat and milk productivity of cattle
 35 breeds (FAOSTAT 2018; Davis et al. 2015).

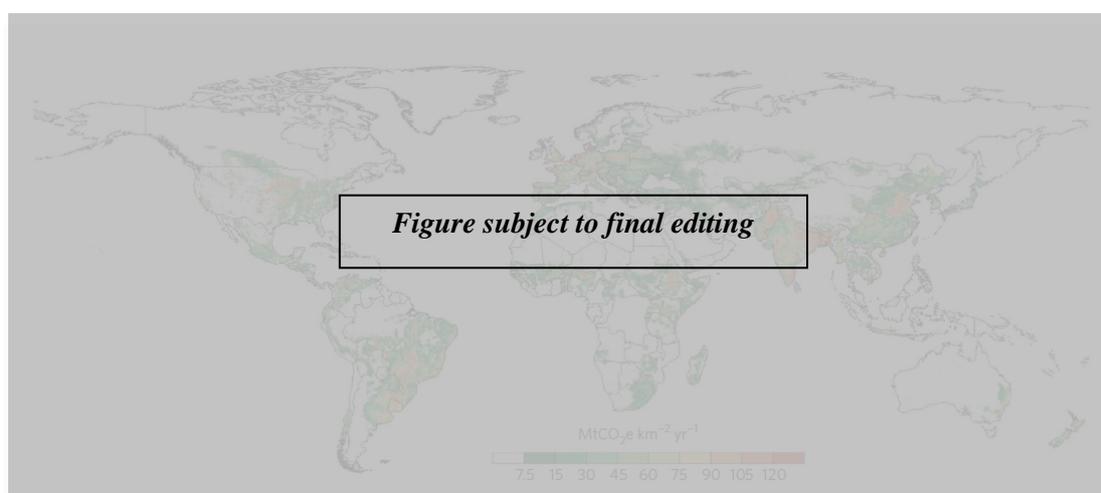
1 Still, products like red meat remain the most inefficient in terms of emissions per kg of protein
2 produced in comparison to milk, pork, eggs and all crop products (IPCC 2014b). Yet, the functional
3 unit used in these measurements is highly relevant and may produce different results (Salou et al.
4 2017). For instance, metrics based on products tend to rate intensive livestock systems as efficient,
5 while metrics based on area or resources used tend to rate extensive system as efficient (Garnett
6 2011). In ruminant dairy systems, less intensified farms show higher emissions if expressed by
7 product, and lower emissions if expressed by Utilizable Agricultural Land (Gutiérrez-Peña et al. 2019;
8 Salvador et al. 2017; Salou et al. 2017).

9 Furthermore, if other variables are used in the analysis of GHG emissions of different ruminant
10 production systems, such as human-edible grains used to feed animals instead of crop waste and
11 pastures of marginal lands, or C sequestration in pasture systems in degraded lands, then the GHG
12 emissions of extensive systems are reduced. Reductions of 26 and 43% have been shown in small
13 ruminants, such as sheep and goats (Gutiérrez-Peña et al. 2019; Salvador et al. 2017; Batalla et al.
14 2015; Petersen et al. 2013). In this regard, depending on what the main challenge is in different
15 regions (e.g., undernourishment, overconsumption, natural resources degradation), different metrics
16 could be used as reference. Other metrics that consider nutrient density have been proposed because
17 they provide potential for addressing both mitigation and health targets (Doran-Browne et al. 2015).

18 Uncertainty in worldwide livestock population numbers remain the main source of variation in total
19 emissions of the livestock sector, while at the animal level, feed intake, diet regime, and nutritional
20 composition are the main sources of variation through their impacts on enteric fermentation and
21 manure N excretion.

22 Increases in economies of scale linked to increased efficiencies and decreased emission intensities
23 may lead to more emissions, rather than less, an observed dynamic referred to by economists as a
24 'rebound effect.' This is because increased efficiency allows production processes to be performed
25 using fewer resources and often at lower cost. This in turn influences consumer behaviour and product
26 use, increasing demand and leading to increased production. In this way, the expected gains from new
27 technologies that increase the efficiency of resource use may be reduced (i.e., increase in the total
28 production of livestock despite increased efficiency of production due to increased demand for meat
29 sold at lower prices). Thus, in order for the livestock sector to provide a contribution to GHG
30 mitigation, reduction in emissions intensities need to be accompanied by appropriate governance and
31 incentive mechanisms to avoid rebound effects, such as limits on total production.

32



33

34 **Figure 5.9 Global GHG emissions from livestock for 1995-2005 (Herrero et al. 2016a)**

35 Variation in estimates of N₂O emissions are due to differing a) climate regimes, b) soil types, and c)
36 N transformation pathways (Charles et al. 2017; Fitton et al. 2017). It was recently suggested that

Subject to Copy-editing

1 N₂O soil emissions linked to livestock through manure applications could be 20%-40% lower than
2 previously estimated in some regions, for instance in Sub-Saharan Africa and Eastern Europe (Gerber
3 et al. 2016) and from smallholder systems in East Africa (Pelster et al. 2017). Herrero et al. (2016a)
4 estimated global livestock enteric methane to range from 1.6–2.7 Gt CO₂-eq, depending on
5 assumptions of body weight and animal diet.

6 7 **5.4.4 Greenhouse gas emissions from aquaculture**

8 Emissions from aquaculture and fisheries may represent some 10% of total agriculture emissions, or
9 about 0.58 Gt CO₂-eq yr⁻¹ (Barange et al. 2018), with two-thirds being non-CO₂ emissions from
10 aquaculture (Hu et al. 2013; Yang et al. 2015) and the rest due to fuel use in fishing vessels. They
11 were not included in Table 5.4 under agriculture emissions, as these estimates are not included in
12 national GHG inventories and global numbers are small as well as uncertain.

13 Methodologies to measure aquaculture emissions are still being developed (Vasanth et al. 2016). N₂O
14 emissions from aquaculture are partly linked to fertiliser use for feed as well as aquatic plant growth,
15 and depend on the temperature of water as well as on fish production (Paudel et al. 2015). Hu et al.
16 (2012) estimated the global N₂O emissions from aquaculture in 2009 to be 0.028 Gt CO₂-eq yr⁻¹, but
17 could increase to 0.114 Gt CO₂-eq yr⁻¹ (that is 5.72% of anthropogenic N₂O–N emissions) by 2030 for
18 an estimated 7.10% annual growth rate of the aquaculture industry. Numbers estimated by Williams
19 and Crutzen (2010) were around 0.036 Gt CO₂-eq yr⁻¹, and suggested that this may rise to more than
20 0.179 Gt CO₂-eq yr⁻¹ within 20 years for an estimated annual growth of 8.7%. (Barange et al. 2018)
21 assessed the contribution of aquaculture to climate change as 0.38 Gt CO₂-eq yr⁻¹ in 2010, around 7%
22 of those from agriculture.

23 CO₂ emissions coming from the processing and transport of feed for fish raised in aquaculture, and
24 also the emissions associated with the manufacturing of floating cultivation devices (e.g., rafts or
25 floating fish-farms), connecting or mooring devices, artificial fishing banks or reefs, and feeding
26 devices (as well as their energy consumption) may be considered within the emissions from the food
27 system. Indeed, most of the GHG emissions from aquaculture are associated with the production of
28 raw feed materials and secondarily, with the transport of raw materials to mills and finished feed to
29 farms (Barange et al. 2018).

30 31 **5.4.5 Greenhouse gas emissions from inputs, processing, storage, and transport**

32 Apart from emissions from agricultural activities within the farm gate, food systems also generate
33 emissions from the pre- and post-production stages in the form of input manufacturing (fertilisers,
34 pesticides, feed production) and processing, storage, refrigeration, retail, waste disposal, food service,
35 and transport. The total contribution of these combined activities outside the farm gate is not well
36 documented. Based on information reported in the AR5 (Fischedick et al. 2014), we estimated their
37 total contribution to be roughly 15% of total anthropogenic emissions (Table 5.4). There is no post-
38 AR5 assessment at the global level in terms of absolute emissions. Rather, several studies have
39 recently investigated how the combined emissions within and outside the farm gate are embedded in
40 food products and thus associated with specific dietary choices (see next section). Below important
41 components of food systems emissions beyond the farm gate are discussed based on recent literature.

42 Refrigerated trucks, trailers, shipping containers, warehouses, and retail displays that are vital parts of
43 food supply chains all require energy and are direct sources of GHG emissions. Upstream emissions
44 in terms of feed and fertiliser manufacture and downstream emissions (transport, refrigeration) in
45 intensive livestock production (dairy, beef, pork) can account for up to 24–32% of total livestock
46 emissions, with the higher fractions corresponding to commodities produced by monogastric animals

1 (Weiss and Leip 2012). The proportion of upstream/downstream emissions fall significantly for less-
2 intensive and more-localised production systems (Mottet et al. 2017a).

3 *Transport and processing.* Recent globalisation of agriculture has promoted industrial agriculture and
4 encouraged value-added processing and more distant transport of agricultural commodities, all
5 leading to increased GHG emissions. Although greenhouse gas-intensive, food transportation plays an
6 important role in food chains: it delivers food from producers to consumers at various distances,
7 particularly to feed people in food-shortage zones from food-surplus zones. (See Section **Error!**
8 **reference source not found.** for assessment of local food production.)

9 To some extent, processing is necessary in order to make food supplies more stable, safe, long-lived,
10 and in some cases, nutritious (FAO 2007). Agricultural production within the farm gate may
11 contribute 80–86% of total food-related emissions in many countries, with emissions from other
12 processes such as processing and transport being small (Vermeulen et al. 2012). However, in net
13 food-importing countries where consumption of processed food is common, emissions from other
14 parts of the food life cycle generated in other locations are much higher (Green et al. 2015).

15 A study conducted by Wakeland et al. (2012) in the US found that the transportation-related carbon
16 footprint varies from a few percent to more than half of the total carbon footprint associated with food
17 production, distribution, and storage. Most of the GHGs emitted from food processing are a result of
18 the use of electricity, natural gas, coal, diesel, gasoline or other energy sources. Cookers, boilers, and
19 furnaces emit carbon dioxide, and wastewater emits methane and nitrous oxide. The most energy-
20 intensive processing is wet milling of maize, which requires 15% of total US food industry energy
21 (Bernstein et al. 2008); processing of sugar and oils also requires large amounts of energy.

22

23 **5.4.6 Greenhouse gas emissions associated with different diets**

24 There is now an extensive literature on the relationship between food products and emissions,
25 although the focus of the studies has been on high-income countries. Godfray et al. (2018) updated
26 Nelson et al. (2016), a previous systematic review of the literature on environmental impacts
27 associated with food, and concluded that higher consumption of animal-based foods was associated
28 with higher estimated environmental impacts, whereas increased consumption of plant-based foods
29 was associated with estimated lower environmental impact. Assessment of individual foods within
30 these broader categories showed that meat – sometimes specified as ruminant meat (mainly beef) –
31 was consistently identified as the single food with the greatest impact on the environment, most often
32 in terms of GHG emissions and/or land use per unit commodity. Similar hierarchies, linked to well-
33 known energy losses along trophic chains, from roots to beef were found in another recent review
34 focussing exclusively on GHG emissions (Clune et al. 2017), and one on life-cycle assessments
35 (Poore and Nemecek 2018). Poore and Nemecek (2018) amassed an extensive database that specifies
36 both the hierarchy of emissions intensities and the variance with the production context (i.e., by
37 country and farming system).

38 The emissions intensities of red meat mean that its production has a disproportionate impact on total
39 emissions (Godfray et al. 2018). For example, in the US 4% of food sold (by weight) is beef, which
40 accounts for 36% of food-related emissions (Heller and Keoleian 2015). Food-related emissions are
41 therefore very sensitive to the amount and type of meat consumed. However, 100 g of beef has twice
42 as much protein as the equivalent in cooked weight of beans, for example, and 2.5 times more iron.
43 One can ingest only about 2.5 kg of food per day and not all food items are as dense in nutrition.

44 There is therefore *robust evidence with high agreement* that the mixture of foods eaten can have a
45 highly significant impact on per capita carbon emissions, driven particularly through the amount of
46 (especially grain-fed) livestock and products.

1 Given the rising costs of malnutrition in all its forms, a legitimate question is often asked: would a
2 diet that promotes health through good nutrition also be one that mitigates GHG emissions? Whilst
3 sustainable diets need not necessarily provide more nutrition, there is certainly significant overlap
4 between those that are healthier (e.g., via eating more plant-based material and less livestock-based
5 material), and eating the appropriate level of calories. In their systematic review, Nelson et al. (2016)
6 conclude that, in general, a dietary pattern that is higher in plant-based foods, such as vegetables,
7 fruits, whole grains, legumes, nuts, and seeds, and lower in animal-based foods is more health-
8 promoting and is associated with lesser environmental impact (GHG emissions and energy, land, and
9 water use) than is the current average “meat-based” diet.

10 Recent FAO projections of food and agriculture to 2050 under alternative scenarios characterised by
11 different degrees of sustainability, provide global-scale evidence that rebalancing diets is key to
12 increasing the overall sustainability of food and agricultural systems world-wide. A 15% reduction of
13 animal products in the diets of high-income countries by 2050 would contribute to containing the need
14 to expand agricultural output due to upward global demographic trends. Not only would GHG
15 emissions and the pressure on land and water be significantly reduced but the potential for low-
16 income countries to increase the intake of animal-based food, with beneficial nutritional outcomes,
17 could be enhanced (FAO 2018a). Given that higher-income countries typically have higher emissions
18 per capita, results are particularly applicable in such places.

19 However, Springmann et al. (2018a) found that there are locally applicable upper bounds to the
20 footprint of diets around the world, and for lower-income countries undergoing a nutrition transition,
21 adopting “Westernised” consumption patterns (over consumption, large amounts of livestock produce,
22 sugar and fat), even if in culturally applicable local contexts, would increase emissions. The global
23 mitigation potential of healthy but low-emissions diets is discussed in detail in Section **Error!**
24 **reference source not found.**

25 In summary, food system emissions are growing globally due to increasing population, income, and
26 demand for animal-sourced products (*high confidence*). Diets are changing on average toward greater
27 consumption of animal-based foods, vegetable oils and sugar/sweeteners (*high confidence*) (see also
28 Chapter 2), with GHG emissions increasing due to greater amounts of animal-based products in diets
29 (*robust evidence, medium agreement*).

30

31 **5.5 Mitigation options, challenges, and opportunities**

32 The IPCC AR5 WG III concluded that mitigation in agriculture, forestry, and land use (AFOLU) is
33 key to limit climate change in the 21st century, in terms of mitigation of non-CO₂ GHGs, which are
34 predominately emitted in AFOLU, as well as in terms of land-based carbon sequestration. Wollenberg
35 et al. (2016) highlighted the need to include agricultural emissions explicitly in national mitigation
36 targets and plans, as a necessary strategy to meet the 2°C goal of the Paris Agreement. This chapter
37 expands on these key findings to document how mitigation in the entire food system, from farm ate to
38 consumer, can contribute to reaching the stated global mitigation goals, but in a context of improved
39 food security and nutrition. To put the range of mitigation potential of food systems in context, it is
40 worth noting that emissions from crop and livestock are expected to increase by 30-40% from present
41 to 2050, under business-as-usual scenarios that include efficiency improvements as well as dietary
42 changes linked to increased income per capita (FAO 2018a; Tubiello et al. 2014). Using current
43 emissions estimates in this chapter and Chapter 2, these increases translate into projected GHG
44 emissions from agriculture of 8-9 Gt CO₂eq yr⁻¹ by 2050 (*medium confidence*).

45 The AR5 ranked mitigation measures from simple mechanisms such as improved crop and livestock
46 management (Smith et al. 2014) to more complex carbon dioxide reduction interventions, such as

1 afforestation, soil carbon storage and biomass energy projects with carbon capture and storage
2 (BECCS). The AR5 WGIII AFOLU chapter (Smith et al. 2014) identified two primary categories of
3 mitigation pathways from the food system:

4 *Supply side*: Emissions from agricultural soils, land use change, land management, and crop and
5 livestock practices can be reduced and terrestrial carbon stocks can be increased by increased
6 production efficiencies and carbon sequestration in soils and biomass, while emissions from energy
7 use at all stages of the food system can be reduced through improvements in energy efficiency and
8 fossil fuel substitution with carbon-free sources including biomass.

9 *Demand side*: GHG emissions could be mitigated by changes in diet, reduction in food loss and waste,
10 and changes in wood consumption for cooking.

11 In this chapter, supply-side mitigation practices include land use change and carbon sequestration in
12 soils and biomass in both crop and livestock systems. Cropping systems practices include improved
13 land and fertiliser management, land restoration, biochar applications, breeding for larger root
14 systems, and bridging yield gaps (Dooley and Stabinsky 2018). Options for mitigation in livestock
15 systems include better manure management, improved grazing land management, and better feeding
16 practices for animals. Agroforestry also is a supply-side mitigation practice. Improving efficiency in
17 supply chains is a supply-side mitigation measure.

18 Demand-side mitigation practices include dietary changes that lead to reduction of GHG emissions
19 from production and changes in land use that sequester carbon. Reduction of food loss and waste can
20 contribute to mitigation of GHGs on both the supply and demand sides. See Section 5.7 and Chapter 7
21 for the enabling conditions needed to ensure that these food system measures would deliver their
22 potential mitigation outcomes.

23

24 **5.5.1 Supply-side mitigation options**

25 The IPCC AR5 identified options for GHG mitigation in agriculture including cropland management,
26 restoration of organic soils, grazing land management and livestock, with a total mitigation potential
27 of 1.5–4.0 GtCO₂-eq yr⁻¹ by 2030 (compared to baseline emissions in the same year), at carbon prices
28 from 20 to 100 USD/tCO₂-eq (Smith et al. 2014). Reductions in GHG emissions intensity (emissions
29 per unit product) from livestock and animal products can also be a means to achieve reductions in
30 absolute emissions in specific contexts and with appropriate governance (*medium confidence*).
31 Agroforestry mitigation practices include rotational woodlots, long-term fallow, and integrated land
32 use.

33 Emissions from food systems can be reduced significantly by the implementation of practices that
34 reduce carbon dioxide, methane, and nitrous oxide emissions from agricultural activities related to the
35 production of crops, livestock, and aquaculture. These include implementation of more sustainable
36 and efficient crop and livestock production practices aimed at reducing the amount of land needed per
37 output (reductions in GHG emissions intensity from livestock and animal production can support
38 reductions in absolute emissions if total production is constrained), bridging yield gaps, implementing
39 better feeding practices for animals and fish in aquaculture, and better manure management (FAO
40 2019a). Practices that promote soil improvements and carbon sequestration can also play an important
41 role. In the South America region, reduction of deforestation, restoration of degraded pasture areas,
42 and adoption of agroforestry and no-till agricultural techniques play a major role in the nation's
43 voluntary commitments to reduce GHG emissions in the country's mitigation activities (Box 5.4).

44 The importance of supply-side mitigation options is that these can be directly applied by food system
45 actors (farmers, processors, retailers, etc.) and can contribute to improved livelihoods and income
46 generation. Recognising and empowering farming system actors with the right incentives and

1 governance systems will be crucial to increasing the adoption rates of effective mitigation practices
2 and to build convincing cases for enabling GHG mitigation (Section 5.7 and Chapter 7).

4 **Box 5.4 Towards sustainable intensification in South America region**

5 Reconciling the increasing global food demand with limited land resources and low environmental
6 impact is a major global challenge (FAO 2018a; Godfray and Garnett 2014; Yao et al. 2017). South
7 America has been a significant contributor of the world's agricultural production growth in the last
8 three decades (OECD and FAO 2015), driven partly by increased export opportunities for specific
9 commodities, mainly soybeans and meat (poultry, beef and pork).

10 Agricultural expansion, however, has driven profound landscape transformations in the region,
11 particularly between the 1970s and early 2000s, contributing to increased deforestation rates and
12 associated GHG emissions. High rates of native vegetation conversion were found in Argentina,
13 Bolivia, Brazil, Colombia, Ecuador, Paraguay and Peru (FAO 2016b; Graesser et al. 2015),
14 threatening ecologically important biomes, such as the Amazon, the savannas (Cerrado, Chacos and
15 Lannos), the Atlantic Rainforest, the Caatinga, and the Yungas. The Amazon biome is a particularly
16 sensitive biome as it provides crucial ecosystem services including biodiversity, hydrological
17 processes (through evapotranspiration, cloud formation, and precipitation), and biogeochemical cycles
18 (including carbon) (Bogaerts et al. 2017; Fearnside 2015; Beuchle et al. 2015; Grecchi et al. 2014;
19 Celentano et al. 2017; Soares-Filho et al. 2014; Nogueira et al. 2018). Further, deforestation
20 associated with commodity exports has not led to inclusive socioeconomic development, but rather
21 has exacerbated social inequality and created more challenging living conditions for lower-income
22 people (Celentano et al. 2017); nor has it avoided increased hunger of local populations in the last few
23 years (FAO 2018b).

24 In the mid-2000s, governments, food industries, NGOs, and international programs joined forces to
25 put in place important initiatives to respond to the growing concerns about the environmental impacts
26 of agricultural expansion in the region (Negra et al. 2014; Finer et al. 2018). Brazil led regional action
27 by launching the Interministerial Plan of Action for Prevention and Control of Deforestation of the
28 Legal Amazon² (PPCDAm), associated with development of a real-time deforestation warning
29 system. Further, Brazil built capacity to respond to alerts by coordinated efforts of ministries, the
30 federal police, the army and public prosecution (Negra et al. 2014; Finer et al. 2018).

31 Other countries in the region have also launched similar strategies, including a zero-deforestation plan
32 in Paraguay in 2004 (Gasparri and de Waroux 2015), and no-deforestation zones in Argentina in 2007
33 (Garcia Collazo et al. 2013). Peru also developed the National System of Monitoring and Control, led
34 by the National Forest Service and Wildlife Authority (SERFOR), to provide information and
35 coordinate response to deforestation events, and Colombia started producing quarterly warning
36 reports on active fronts of deforestation in the country (Finer et al. 2018).

37 Engagement of the food industry and NGOs, particularly through the Soy Moratorium (from 2006)
38 and Beef Moratorium (from 2009) also contributed effectively to keep deforestation at low historical
39 rates in the regions where they were implemented (Nepstad et al. 2014; Gibbs et al. 2015). In 2012,
40 Brazil also created the national land registry system (SICAR), a georeferenced database, which allows
41 monitoring of farms' environmental liability in order to grant access to rural credit. Besides the

² FOOTNOTE: The Legal Amazon is a Brazilian region of 501.6 Mha (ca. 59% of the Brazilian territory) that contains all the Amazon but also 40% of the Cerrado and 40% of the Pantanal biomes, with a total population of 25.47 million inhabitants.

1 governmental schemes, funding agencies and the Amazon Fund provide financial resources to assist
2 smallholder farmers to comply with environmental regulations (Jung et al. 2017).

3 Nevertheless, Azevedo et al. (2017) argue that the full potential of these financial incentives has not
4 been achieved, due to weak enforcement mechanisms and limited supporting public policies.
5 Agricultural expansion and intensification have complex interactions with deforestation. While
6 mechanisms have been implemented in the region to protect native forests and ecosystems, control of
7 deforestation rates require stronger governance of natural resources (Ceddia et al. 2013; Oliveira and
8 Hecht 2016), including monitoring programs to evaluate fully the results of land use policies in the
9 region.

10 Public and private sector actions resulted in a reduction of the Brazilian legal Amazon deforestation
11 rate from 2.78 Mha yr⁻¹ in 2004, to about 0.75 Mha yr⁻¹ (ca. 0.15%) in 2009 (INPE 2015), oscillating
12 from 0.46 Mha and 0.79 Mha (2016) since then (INPE 2018; Boucher and Chi 2018). The
13 governmental forest protection scheme was also expanded to other biomes. As a result, the Brazilian
14 Cerrado deforestation was effectively reduced from 2.9 Mha yr⁻¹ in 2004 to an average of 0.71 Mha
15 yr⁻¹ in 2016-2017 (INPE 2018).

16 Overall, deforestation rates in South America have declined significantly, with current deforestation
17 rates being about half of rates in the early 2000s (FAOSTAT 2018). However, inconsistent
18 conservation policies across countries (Gibbs et al. 2015) and recent hiccups (Curtis et al. 2018)
19 indicate that deforestation control still requires stronger reinforcement mechanisms (Tollefson 2018).
20 Further, there are important spill-over effects that need coordinated international governance. Curtis et
21 al. (2018) and Dou et al. (2018) point out that, although the Amazon deforestation rate decreased in
22 Brazil, it has increased in other regions, particularly in Southern Asia, and in other countries in South
23 America, resulting in nearly constant deforestation rates worldwide.

24 Despite the reduced expansion rates into forest land, agricultural production continues to rise steadily
25 in South America, relying on increasing productivity and substitution of extensive pastureland by
26 crops. The average soybean and maize productivity in the region increased from 1.8 and 2.0 t ha⁻¹ in
27 1990 to 3.0 and 5.0 t ha⁻¹, respectively, in 2015 (FAOSTAT 2018). Yet, higher crop productivity was
28 not enough to meet growing demand for cereals and oilseeds and cultivation continued to expand,
29 mainly on grasslands (Richards 2015). The reconciliation of this expansion with higher demand for
30 meat and dairy products was carried out through the intensification of livestock systems (Martha et al.
31 2012). Nevertheless, direct and indirect deforestation still occurs, and recently deforestation rates have
32 increased (INPE 2018), albeit they remain far smaller than observed in the 2000-2010 period.

33 The effort towards sustainable intensification has also been incorporated in agricultural policies. In
34 Brazil, for instance, the reduction of deforestation, the restoration of degraded pasture areas, the
35 adoption of integrated agroforestry systems³ and no-till agricultural techniques play a major role in
36 the nation's voluntary commitments to reduce GHG emissions in the country's NAMAs (Mozzer
37 2011) and NDCs (Silva Oliveira et al. 2017; Rochedo et al. 2018). Such commitment under the
38 UNFCCC is operationalised through the Low Carbon Agriculture Plan (ABC)⁴, which is based on low
39 interest credit for investment in sustainable agricultural technologies (Mozzer 2011). Direct pasture
40 restoration and integrated systems reduce area requirements (Strassburg et al. 2014), and increase
41 organic matter (Gil et al. 2015; Bungenstab 2012; Maia et al. 2009), contributing to overall life cycle

³ FOOTNOTE: Integrated agroforestry systems are agricultural systems that strategically integrate two or more components among crops, livestock and forestry. The activities can be in consortium, succession or rotation in order to achieve overall synergy.

⁴ FOOTNOTE: ABC - *Agricultura de Baixo Carbono* in Portuguese.

1 emissions reduction (Cardoso et al. 2016; de Oliveira Silva et al. 2016). Also, increased adoption of
2 supplementation and feedlots, often based on agro-industrial co-products and agricultural crop
3 residues are central to improve productivity and increase climate resilience of livestock systems
4 (Mottet et al. 2017a; van Zanten et al. 2018).

5 Despite providing clear environmental and socio-economic co-benefits, including improved resource
6 productivity, socio-environmental sustainability and higher economic competitiveness,
7 implementation of the Brazilian Low Carbon Agriculture Plan is behind schedule (Köberle et al.
8 2016). Structural inefficiencies related to the allocation and distribution of resources need to be
9 addressed to put the plan on track to meet its emissions reduction targets. Monitoring and verification
10 are fundamental tools to guarantee the successful implementation of the plan.

11 Overall, historical data and projections show that South America is one of the regions of the world
12 with the highest potential to increase crop and livestock production in the coming decades in a
13 sustainable manner (Cohn et al. 2014), increasing food supply to more densely populated regions in
14 Asia, Middle East and Europe. However, a great and coordinated effort is required from governments,
15 industry, traders, scientists and the international community to improve planning, monitoring and
16 innovation to guarantee sustainable intensification of its agricultural systems, contribution to GHG
17 mitigation, and conservation of the surrounding environment (Negra et al. 2014; Curtis et al. 2018;
18 Lambin et al. 2018).

19 20 **5.5.1.1 Greenhouse gas mitigation in croplands and soils**

21 The mitigation potential of agricultural soils, cropland and grazing land management has been the
22 subject of much research and was thoroughly summarised in the AR5 (Smith et al. 2014) (See also
23 Chapter 2 Section 2.5.1 and Chapter 6 Section 6.3.1). Key mitigation pathways are related to practices
24 reducing nitrous oxide emissions from fertiliser applications, reducing methane emissions from paddy
25 rice, reducing both gases through livestock manure management and applications, and sequestering
26 carbon or reducing its losses, with practices for improving grassland and cropland management
27 identified as the largest mitigation opportunities. Better monitoring reporting and verification (MRV)
28 systems are currently needed for reducing uncertainties and better quantifying the actual mitigation
29 outcomes of these activities.

30 New work since AR5 has focused on identifying pathways for the reductions of GHG emissions from
31 agriculture to help meet Paris Agreement goals (Paustian et al. 2016; Wollenberg et al. 2016). Altieri
32 and Nicholls (2017) have characterised mitigation potentials from traditional agriculture. Zomer et al.
33 (2017) have updated previous estimates of global carbon sequestration potential in cropland soils.
34 Mayer et al. (2018) converted soil carbon sequestration potential through agricultural land
35 management into avoided temperature reductions. Fujisaki et al. (2018) identify drivers to increase
36 soil organic carbon in tropical soils. For discussion of integrated practices such as sustainable
37 intensification, conservation agriculture and agroecology, see Section 5.6.4.

38 Paustian et al. (2016) developed a decision-tree for facilitating implementation of mitigation practices
39 on cropland and described the features of key practices. They observed that most individual mitigation
40 practices will have a small effect per unit of land, and hence they need to be combined and applied at
41 large scales for their impact to be significant. Examples included aggregation of cropland practices
42 (e.g., organic amendments, improved crop rotations and nutrient management and reduced tillage) and
43 grazing land practices (e.g., grazing management, nutrient and fire management and species
44 introduction) that could increase net soil C stocks while reducing emissions of N₂O and CH₄.
45 However, it is well-known that the portion of projected mitigation from soil C stock increase (about
46 90% of the total technical potential) is impermanent, i.e., it would be effective for only 20–30 years
47 due to saturation of the soil capacity to sequester carbon, whereas non-CO₂ emission reductions could

1 continue indefinitely. “Technical potential” is the maximum amount of GHG mitigation achievable
2 through technology diffusion.

3 Biochar application and management towards enhanced root systems are mitigation options that have
4 been highlighted in recent literature (Dooley and Stabinsky 2018; Hawken 2017; Paustian et al. 2016;
5 Woolf et al. 2010; Lenton 2010).

6

7 **5.5.1.2 Greenhouse gas mitigation in livestock systems**

8 The technical options for mitigating GHG emissions in the livestock sector have been the subject of
9 recent reviews (Mottet et al., 2017b; Hristov et al. 2013a,b; Smithers 2015; Herrero et al. 2016a;
10 Rivera-Ferre et al. 2016b) (Figure 5.11). They can be classified as either targeting reductions in
11 enteric methane; reductions in nitrous oxide through manure management; sequestering carbon in
12 pastures; implementation of best animal husbandry and management practices, which would have an
13 effect on most GHG; and land use practices that also help sequester carbon. Excluding land use
14 practices, these options have a technical mitigation potential ranging 0.2-2.4 GtCO₂-eq yr⁻¹ (Herrero et
15 al. 2016a; FAO 2007). See also Chapters 2 and 6 in this report.

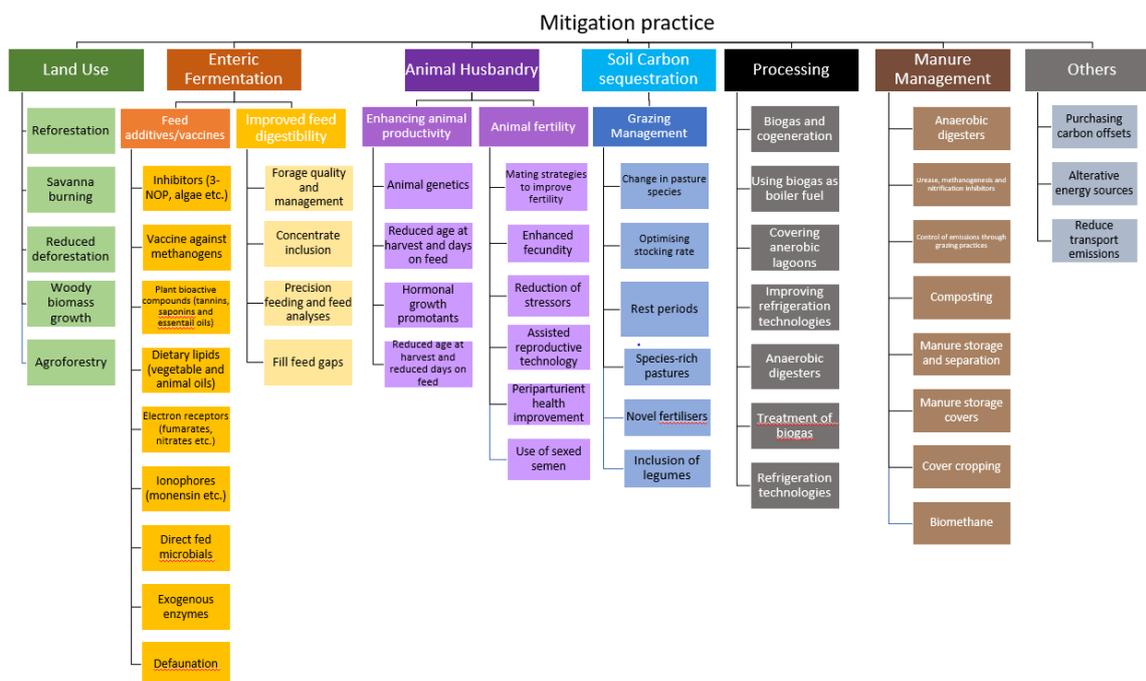
16 The opportunities for carbon sequestration in grasslands and rangelands may be significant (Conant
17 2010), for instance, through changes in grazing intensity or manure recycling aimed at maintaining
18 grassland productivity (Hirata et al. 2013). Recent studies have questioned the economic potential of
19 such practices, i.e., whether they could implement at scale for economic gain (Garnett et al. 2017;
20 Herrero et al. 2016a; Henderson et al. 2015). For instance, Henderson et al. (2015) found economic
21 potentials below 200 MtCO₂-eq yr⁻¹. Carbon sequestration can occur in situations where grasslands
22 are highly degraded (Garnett 2016). Carbon sequestration linked to livestock management could thus
23 be considered as a co-benefit of well-managed grasslands, as well as a mitigation practice.

24 Different production systems will require different strategies, including the assessment of impacts on
25 food security, and this has been the subject of significant research (e.g., Rivera-Ferre et al. 2016b).
26 Livestock systems are heterogeneous in terms of their agro-ecological orientation (arid, humid or
27 temperate/highland locations), livestock species (cattle, sheep, goats, pigs, poultry and others),
28 structure (grazing only, mixed-crop-livestock systems, industrial systems, feedlots and others), level
29 of intensification, and resource endowment (Robinson 2011).

30 The implementation of strategies presented in **Error! Reference source not found.** builds on this
31 differentiation, providing more depth compared to the previous AR5 analysis. Manure management
32 strategies are more applicable in confined systems, where manure can be easily collected, such as in
33 pigs and poultry systems or in smallholder mixed crop–livestock systems. More intensive systems,
34 with strong market orientation, such as dairy in the US, can implement a range of sophisticated
35 practices like feed additives and vaccines, while many market-oriented dairy systems in tropical
36 regions can improve feed digestibility by improving forage quality and adding larger quantities of
37 concentrate to the rations. Many of these strategies can be implemented as packages in different
38 systems, thus maximising the synergies between different options (Mottet et al. 2017b).

39 See the Supplementary Material Section SM5.5 for detailed description of livestock mitigation
40 strategies; synergies and trade-offs with other mitigation and adaptation options are discussed in
41 Section 5.6.

42



1
2 **Figure 5.10 Technical supply-side mitigation practices in the livestock sector (adapted from Hristov et al.**
3 **2013b; Herrero et al. 2016b; Smith et al. 2014)**

4 **5.5.1.3 Greenhouse gas mitigation in agroforestry**

5 Agroforestry can curb GHG emissions of CO₂, CH₄, and N₂O in agricultural systems in both
6 developed and developing countries (See Glossary for definition) (see Chapter 2, Section 2.5.1 and
7 Figure 2.24). Soil carbon sequestration, together with biological N fixation, improved land health and
8 underlying ecosystem services may be enhanced through agricultural lands management practices
9 used by large-scale and smallholder farmers, such as incorporation of trees within farms or in hedges
10 (manure addition, green manures, cover crops, etc.), whilst promoting greater soil organic matter and
11 nutrients (and thus soil organic carbon) content and improve soil structure (Mbow et al. 2014b) (Table
12 5.5). The tree cover increases the microbial activity of the soil and increases the productivity of the
13 grass under cover. CO₂ emissions are furthermore lessened indirectly, through lower rates of erosion
14 due to better soil structure and more plant cover in diversified farming systems than in monocultures.
15 There is great potential for increasing above ground and soil C stocks, reducing soil erosion and
16 degradation, and mitigating GHG emissions.

17 These practices can improve food security through increases in productivity and stability since they
18 contribute to increased soil quality and water-holding capacity. Agroforestry provides economic,
19 ecological, and social stability through diversification of species and products. On the other hand,
20 trade-offs are possible when cropland is taken out of production mainly as a mitigation strategy.

21
22 **Table 5.5 Carbon sequestration potential for agroforestry (Mbow et al. 2014b)**

| Source | Carbon sequestration (tCO ₂ km ⁻² yr ⁻¹) (range) | C stock (tCO ₂ km ⁻²) (range) | Max rotation period (years) |
|--|---|---|--------------------------------|
| Dominant parklands | 183 (73–293) | 12,257 (2,091–25,983) | 50 |
| Rotational woodlots^a | 1,431 (807–2,128) | 6,789 (4,257–9,358) | 5 |

| | | | |
|--|----------------------|---------------------------|----|
| Tree planting-windrows-home gardens | 220.2 (146–293) | 6,973 (-) | 25 |
| Long term fallows, regrowth of woodlands in abandoned farms^b | 822 (80–2,128) | 5,761 (-) | 25 |
| Integrated land use | 1,145 (367–2,458) | 28,589 (4,404–83,676) | 50 |
| Soil carbon | 330 (91–587) | 33,286 (4,771–110,100) | - |

^a May be classified as forestry on forest land, depending on the spatial and temporal characteristics of these activities.

^b This is potentially not agro-forestry, but forestry following abandonment of agricultural land.

Meta-analyses have been done on carbon budgets in agroforestry systems (Zomer et al. 2016; Chatterjee et al. 2018). In a review of 42 studies, (Ramachandran Nair et al. 2009) estimated C sequestration potentials of differing agroforestry systems. These include sequestration rates in ranging from 954 (semi-arid); to 1,431 (temperate); 2,238 (sub-humid) and 3,670 tCO₂ km⁻² yr⁻¹ (humid). The global technical potential for agroforestry is 0.1–5.7 Gt CO₂e yr⁻¹ (Griscom et al. 2017; Zomer et al. 2016; Dickie et al. 2014) (see Chapter 2, Section 2.5.1). Agroforestry-based carbon sequestration can be used to offset N₂O and CO₂ emissions from soils and increase methane sink strength compared to annual cropping systems (Rosenstock et al. 2014).

Agroforestry systems with perennial crops, such as coffee and cacao, may be more important carbon sinks than those that combine trees with annual crops. Brandt et al. (2018) showed that farms in semi-arid region (300–600 mm precipitation) were increasing in tree cover due to natural regeneration and that the increased application of agroforestry systems were supporting production and reducing GHG emissions.

5.5.1.4 Integrated approaches to crop and livestock mitigation

Livestock mitigation in a circular economy. Novel technologies for increasing the integration of components in the food system are being devised to reduce GHG emissions. These include strategies that help decoupling livestock from land use. Work by van Zanten et al. (2018) shows that 7–23 g of animal protein per capita per day could be produced without livestock competing for vital arable land. This would imply a contraction of in the land area utilised by the livestock sector, but also a more efficient use of resources, and would lead to land sparing and overall emissions reductions.

Pikaar et al. (2018) demonstrated the technical feasibility of producing microbial protein as a feedstuff from sewage that could replace use of feed crops such as soybean. The technical potential of this novel practice could replace 10–19% of the feed protein required, and would reduce cropland demand and associated emissions by 6-7%. These practices are, however, not economically feasible nor easily upscalable in most systems. Nonetheless, significant progress in Japan and South Korea in the reduction and use of food waste to increase efficiencies in livestock food chains has been achieved, indicating a possible pathway to progress elsewhere (FAO 2017; zu Ermgassen et al. 2016). Better understanding of biomass and food and feed wastes, value chains, and identification of mechanisms for reducing the transport and processing costs of these materials is required to facilitate larger-scale implementation.

Waste streams into energy. Waste streams from manure and food waste can be used for energy generation and thus reduction in overall GHG emissions in terms of recovered methane (for instance

1 through anaerobic digestion) production (De Clercq et al. 2016) or for the production of microbial
2 protein (Pikaar et al. 2018). Second-generation biorefineries, once the underlying technology is
3 improved, may enable the generation of hydro-carbon from agricultural residues, grass, and woody
4 biomass in ways that do not compete with food and can generate, along with biofuel, high-value
5 products such as plastics (Nguyen et al. 2017). Second-generation energy biomass from residues may
6 constitute a complementary income source for farmers that can increase their incentive to produce.
7 Technologies include CHP (combined heat and power) or gas turbines, and fuel types such as bio-
8 diesel, bio-pyrolysis (i.e., high temperature chemical transformation of organic material in the absence
9 of oxygen), torrefaction of biomass, production of cellulosic bio-ethanol and of bio-alcohols produced
10 by other means than fermentation, and the production of methane by anaerobic fermentation. (Nguyen
11 et al. 2017).

12 *Technology for reducing fossil fuel inputs.* Besides biomass and bioenergy, other forms of renewable
13 energy substitution for fossil fuels (e.g., wind, solar, geothermal, hydro) are already being applied on
14 farms and throughout the supply chain. Energy efficiency measures are being developed for
15 refrigeration, conservation tillage, precision farming (e.g., fertiliser and chemical application and
16 precision irrigation.

17 *Novel technologies.* Measures that can reduce livestock emissions given continued research and
18 development include methane and nitrification inhibitors, methane vaccines, targeted breeding of
19 lower-emitting animals, and genetically modified grasses with higher sugar content. New strategies to
20 reduce methanogenesis include supplementing animal diets with antimethanogenic agents (e.g., 3-
21 NOP, algae, chemical inhibitors such as chloroform) or supplementing with electron acceptors (e.g.,
22 nitrate) or dietary lipids. These could potentially contribute, once economically feasible at scale, to
23 significant reductions of methane emissions from ruminant livestock. A well-tested compound is 3-
24 nitrooxypropanol (3-NOP), which was shown to decrease methane by up to 40% when incorporated in
25 diets for ruminants (Hristov et al. 2015).

26 Whilst these strategies may become very effective at reducing methane, they can be expensive and
27 also impact on animal performance and/or welfare (Llonch et al. 2017). The use of novel fertilisers
28 and/or plant species that secrete biological nitrification inhibitors also have the potential to
29 significantly reduce N₂O emissions from agricultural soils (Subbarao et al. 2009; Rose et al. 2018).

30 *Economic mitigation potentials of crop and livestock sectors.* Despite the large technical mitigation
31 potential of the agriculture sector in terms of crop and livestock activities, its economic potential is
32 relatively small in the short term (2030) and at modest carbon prices (less than USD 20 tC⁻¹). For crop
33 and soil management practices, it is estimated that 1.0–1.5 GtCO₂-eq yr⁻¹ could be a feasible
34 mitigation target at a carbon price of USD 20/tonne of carbon (Frank et al. 2018, 2017; Griscorn et al.
35 2016; Smith et al. 2013; Wollenberg et al. 2016). For the livestock sector, these estimates range from
36 0.12–0.25 GtCO₂-eq yr⁻¹ at similar carbon prices (Herrero et al. 2016c; Henderson et al. 2017). But
37 care is needed in comparing crop and livestock economic mitigation potentials due to differing
38 assumptions.

39 Frank et al. (2018) recently estimated that the economic mitigation potential of non-CO₂ emissions
40 from agriculture and livestock to 2030 could be up to four times higher than indicated in the AR5, if
41 structural options such as switching livestock species from ruminants to monogastrics, or allowing for
42 flexibility to relocate production to more efficient regions were implemented, at the same time as the
43 technical options such as those described above. At higher carbon prices (i.e., at about USD 100tC⁻¹),
44 they found a mitigation potential of supply-side measures of 2.6 GtCO₂-eq yr⁻¹.

45 In this scenario, technical options would account for 38% of the abatement, while another 38% would
46 be obtained through structural changes, and a further 24% would be obtained through shifts in
47 consumption caused by food price increases. Key to the achievement of this mitigation potential lay in

1 the livestock sector, as reductions in livestock consumption, structural changes and implementation of
2 technologies in the sector had some of the highest impacts. Regions with the highest mitigation
3 potentials were Latin America, China and Sub-Saharan Africa. The large-scale implementability of
4 such proposed sweeping changes in livestock types and production systems is likely very limited as
5 well as constrained by long-established socio-economic, traditional and cultural habits, requiring
6 significant incentives to generate change.

7 In summary, supply-side practices can contribute to climate change mitigation by reducing crop and
8 livestock emissions, sequestering carbon in soils and biomass, and by decreasing emissions intensity
9 within sustainable production systems (*high confidence*). The AR5 estimated the total economic
10 mitigation potential of crop and livestock activities as 1.5–4.0 GtCO₂-eq yr⁻¹ by 2030 at prices ranging
11 from 20-100 USD/tCO₂eq (*high confidence*). Options with large potential for GHG mitigation in
12 cropping systems include soil carbon sequestration (at decreasing rates over time), reductions in N₂O
13 emissions from fertilisers, reductions in CH₄ emissions from paddy rice, and bridging of yield gaps.
14 Options with large potential for mitigation in livestock systems include better grazing land
15 management, with increased net primary production and soil carbon stocks, improved manure
16 management, and higher-quality feed. Reductions in GHG emissions intensity (emissions per unit
17 product) from livestock can support reductions in absolute emissions, provided appropriate
18 governance structures to limit total production are implemented at the same time (*medium*
19 *confidence*).

21 **5.5.1.5 Greenhouse gas mitigation in aquaculture**

22 Barange et al. (2018) provide a synthesis of effective options for GHG emissions reduction in
23 aquaculture including reduction of emissions from production of feed material, replacement of fish-
24 based feed ingredients with crop-based ingredients; reduction of emissions from feed mill energy use,
25 improvement of feed conversion rates, improvement of input use efficiency, shift of energy supply
26 (from high-carbon fossil fuels to low-carbon fossil fuels or renewables), and improvement of fish
27 health. Conversion of 25% of total aquaculture area to integrated aquaculture-agriculture ponds
28 (greening aquaculture) has the potential to sequester 95.4 millions tonnes carbon per year (Ahmed et
29 al. 2017).

30 Proposed mitigation in aquaculture includes avoided deforestation. By halting annual mangrove
31 deforestation in Indonesia, associated total emissions would be reduced 10-31% of estimated annual
32 emissions from land use sector at present (Murdiyarso et al. 2015). Globally, 25% mangrove
33 regeneration could sequester 0.54–0.65 millions tonnes carbon per year (Ahmed et al. 2017) of which
34 0.17-0.21 could be through integrated or organic shrimp culture (Ahmed et al. 2018).

36 **5.5.1.6 Cellular agriculture**

37 The technology for growing muscle tissue in culture from animal stem cells to produce meat, i.e.,
38 “cultured” or “synthetic” or “in vitro” or “hydroponic” meat could in theory be constructed with
39 different characteristics and be produced faster and more efficiently than traditional meat (Kadim et
40 al. 2015). Cultured meat (CM) is part of so-called cellular agriculture, which includes production of
41 milk, egg white and leather from industrial cell cultivation (Stephens et al. 2018). CM is produced
42 from muscle cells extracted from living animals, isolation of adult skeletal muscle stem cells
43 (myosatellite cells), placement in a culture medium which allow their differentiation into myoblasts
44 and then, through another medium, generation of myocytes which coalesce into myotubes and grow
45 into strands in a stirred-tank bioreactor (Mattick et al. 2015). Current technology enables the creation
46 of beef hamburgers, nuggets, steak chips or similar products from meat of other animals, including
47 wild species, although production currently is far from being economically feasible. Nonetheless, by

1 allowing bioengineering from the manipulation of the stem cells and nutritive culture, CM allows for
2 reduction of harmful fatty acids, with advantages such as reduced GHG emissions, mostly indirectly
3 through reduced land use (Bhat et al. 2015; Kumar et al. 2017b).

4 Tuomisto and de Mattos (2011) made optimistic technological assumptions, relying on cyanobacteria
5 hydrolysate nutrient source, and produced the lowest estimates on energy and land use. Tuomisto and
6 de Mattos (2011) conducted a lifecycle assessment that indicates that cultured meat could have less
7 than 60% of energy use and 1% of land use of beef production and it would have lower GHG
8 emissions than pork and poultry as well. Newer estimates (Alexander et al. 2017; Mattick et al. 2015)
9 indicate a trade-off between industrial energy consumption and agricultural land requirements of
10 conventional and cultured meat and possibly higher GWP than pork or poultry due to higher energy
11 use. The change in proportion of CO₂ vs CH₄ could have important implications in climate change
12 projections and, depending on decarbonisation of the energy sources and climate change targets,
13 cultured meat may be even more detrimental than exclusive beef production (Lynch 2019).

14 Overall, as argued by Stephens et al. (2018), cultured meat is an “as-yet undefined ontological
15 object” and, although marketing targets people who appreciate meat but are concerned with animal
16 welfare and environmental impacts, its market is largely unknown (Bhat et al. 2015; Slade 2018). In
17 this context it will face the competition of imitation meat (meat analogues from vegetal protein) and
18 insect-derived products, which have been evaluated as more environmentally friendly (Alexander et
19 al. 2017) and it may be considered as being an option for a limited resource world, rather than a
20 mainstream solution. Besides, as commercial production process is still largely undefined, its actual
21 contribution to climate change mitigation and food security is largely uncertain and challenges are not
22 negligible. Finally, it is important to understand the systemic nature of these challenges and evaluate
23 their social impacts on rural populations due to transforming animal agriculture into an industrialised
24 activity and its possible rebound effects on food security, which are still understudied in the literature.

25 Studies are needed to improve quantification of mitigation options for supply chain activities.

27 **5.5.2 Demand-side mitigation options**

28 Although population growth is one of the drivers of global food demand and the resulting
29 environmental burden, demand-side management of the food system could be one of the solutions to
30 curb climate change. Avoiding food waste during consumption, reducing over-consumption, and
31 changing dietary preferences can contribute significantly to provide healthy diets for all, as well as
32 reduce the environmental footprint of the food system. The number of studies addressing this issue
33 have increased in the last few years (see also Chapter 2). (See Section 5.6 for synergies and trade-offs
34 with health and Section 5.7 for discussion of Just Transitions).

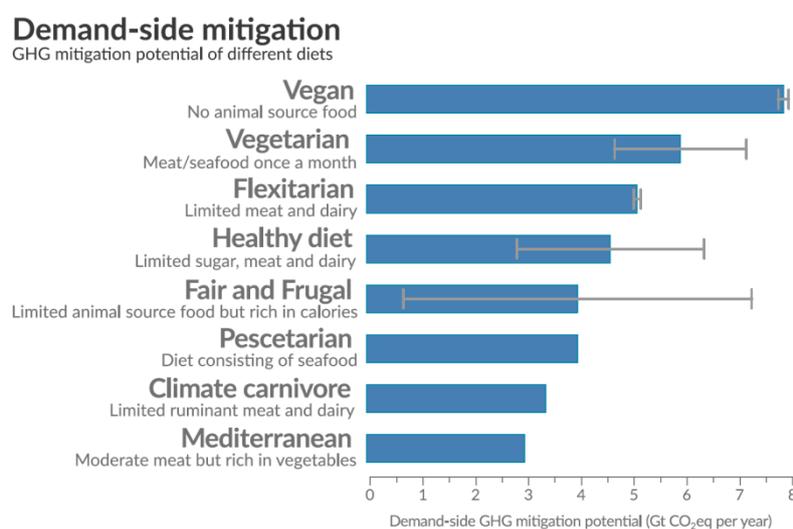
36 **5.5.2.1 Mitigation potential of different diets**

37 A systematic review found that higher consumption of animal-based foods was associated with higher
38 estimated environmental impact, whereas increased consumption of plant-based foods was associated
39 with an estimated lower environmental impact (Nelson et al. 2016). Assessment of individual foods
40 within these broader categories showed that meat – especially ruminant meat (beef and lamb) – was
41 consistently identified as the single food with the greatest impact on the environment, on a global
42 basis, most often in terms of GHG emissions and/or land use.

43 Figure 5.12 shows the technical mitigation potentials of some scenarios of alternative diets examined
44 in the literature. Stehfest et al. (2009) were among the first to examine these questions. They found
45 that under the most extreme scenario, where no animal products are consumed at all, adequate food
46 production in 2050 could be achieved on less land than is currently used, allowing considerable forest

1 regeneration, and reducing land-based greenhouse gas emissions to one third of the reference
 2 “business-as-usual” case for 2050, a reduction of 7.8 Gt CO₂-eq yr⁻¹. Springmann et al. (2016b)
 3 recently estimated similar emissions reduction potential of 8 Gt CO₂-eq yr⁻¹ from a vegan diet without
 4 animal-sourced foods. This defines the upper bound of the technical mitigation potential of demand
 5 side measures.

6 Herrero et al. (2016a) reviewed available options, with a specific focus on livestock products,
 7 assessing technical mitigation potential across a range of scenarios, including “No animal products”,
 8 “No meat”, “No ruminant meat”, and “Healthy diet” (reduced meat consumption). With regard to
 9 ‘credible low-meat diets,’ where reduction in animal protein intake was compensated by higher intake
 10 of pulses, emissions reductions by 2050 could be in the 4.3–6.4 Gt CO₂-eq yr⁻¹, compared to a
 11 business-as-usual scenario. Of this technical potential, 1–2 GtCO₂-eq yr⁻¹ come from reductions of
 12 mostly non-CO₂ GHG within the farm gate, while the remainder was linked to carbon sequestration
 13 on agricultural lands no longer needed for livestock production. When the transition to a low-meat
 14 diet reduces the agricultural area required, land is abandoned and the re-growing vegetation can take
 15 up carbon until a new equilibrium is reached. This is known as the land-sparing effect.



26 **Figure 5.12 Technical mitigation potential of changing diets by 2050 according to a range of scenarios**
 27 **examined in the literature. Estimates are technical potential only, and include additional effects of carbon**
 28 **sequestration from land-sparing. Data without error bars are from one study only.**

29 All diets need to provide a full complement of nutritional quality, including micronutrients (FAO et al. 2018)

30 Vegan: Completely plant-based (Springmann et al. 2016; Stehfest et al. 2009)

31 Vegetarian: Grains, vegetables, fruits, sugars, oils, eggs and dairy, and generally at most one serving per month
 32 of meat or seafood (Springmann et al. 2016; Tilman and Clark 2014; Stehfest et al. 2009)

33 Flexitarian: 75% of meat and dairy replaced by cereals and pulses; at least 500 g per day fruits and vegetables;
 34 at least 100 g per day of plant-based protein sources; modest amounts of animal-based proteins and limited
 35 amounts of red meat (one portion per week), refined sugar (less than 5% of total energy), vegetable oils high in
 36 saturated fat, and starchy foods with relatively high glycaemic index (Springmann et al. 2018; Hedenus et al.
 37 2014)

38 Healthy diet: Based on global dietary guidelines for consumption of red meat, sugar, fruits and vegetables, and
 39 total energy intake (Springmann et al. 2018; Bajželj et al. 2014)

40 Fair and Frugal: Global daily per-capita calorie intake of 2800 kcal/cap/day (11.7 MJ/cap/day), paired with
 41 relatively low level of animal products (Smith et al. 2013)

42 Pescetarian: Vegetarian diet that includes seafood (Tilman and Clark 2014)

43 Climate carnivore: 75% of ruminant meat and dairy replaced by other meat (Hedenus et al. 2014)

44 Mediterranean: Vegetables, fruits, grains, sugars, oils, eggs, dairy, seafood, moderate amounts of poultry, pork,
 45 lamb and beef (Tilman and Clark 2014)

1 Other studies have found similar results for potential mitigation linked to diets. For instance, Smith et
2 al. (2013) analysed a dietary change scenario that assumed a convergence towards a global daily per-
3 capita calorie intake of 2800 kcal cap⁻¹ day⁻¹ (11.7 MJ cap⁻¹ day⁻¹), paired with a relatively low level of
4 animal product supply, estimated technical mitigation potential in the range 0.7–7.3 Gt CO₂-eq yr⁻¹ for
5 additional variants including low or high-yielding bioenergy, 4.6 Gt CO₂-eq yr⁻¹ if spare land is
6 afforested.

7 Bajželj et al. (2014) developed different scenarios of farm systems change, waste management, and
8 dietary change on GHG emissions coupled to land use. Their dietary scenarios were based on target
9 kilocalorie consumption levels and reductions in animal product consumption. Their scenarios were
10 “Healthy Diet”; Healthy Diet with 2500 kcal cap⁻¹ day⁻¹ in 2050; corresponded to technical mitigation
11 potentials in the range 5.8 and 6.4 Gt CO₂-eq yr⁻¹.

12 Hedenus et al. (2014) explored further dietary variants based on the type of livestock product.
13 “Climate Carnivore”, in which 75% of the baseline-consumption of ruminant meat and dairy was
14 replaced by pork and poultry meat, and “Flexitarian”, in which 75% of the baseline-consumption of
15 meat and dairy was replaced by pulses and cereal products. Their estimates of technical mitigation
16 potentials by 2050 ranged 3.4– 5.2 Gt CO₂-eq yr⁻¹, the high end achieved under the Flexitarian.
17 Finally, Tilman and Clark (2014) used stylised diets as variants that included “Pescetarian”,
18 “Mediterranean”, “Vegetarian”, compared to a reference diet, and estimated technical mitigation
19 potentials within the farm gate of 1.2–2.3 Gt CO₂-eq yr⁻¹, with additional mitigation from carbon
20 sequestration on spared land ranging 1.8–2.4 Gt CO₂-eq yr⁻¹.

21 Studies have defined dietary mitigation potential as, for example, 20kg per person per week CO₂-eq
22 for Mediterranean diet, vs 13kg per person per week CO₂-eq for vegan (Castañé and Antón 2017).
23 Rosi et al. (2017) developed seven-day diets in Italy for about 150 people defined as omnivore 4.0 ±
24 1.0; ovo-lacto-veggie 2.6 ± 0.6; and vegan 2.3 ± 0.5 kg CO₂-eq per capita per day.

25 Importantly, many more studies that compute the economic and calorie costs of these scenarios are
26 needed. Herrero et al. (2016a) estimated that once considerations of economic and calorie costs of
27 their diet-based solutions were included, the technical range of 4.3–6.4 Gt CO₂-eq yr⁻¹ in 2050 was
28 reduced to 1.8–3.4 Gt CO₂-eq yr⁻¹ when implementing a GHG tax ranging from 20–100 USD/tCO₂.
29 While caloric costs were low below 20 USD/tCO₂, they ranged from 27–190 kcal cap⁻¹ day⁻¹ under
30 the higher economic potential, thus indicating possible negative trade-offs with food security.

31 In summary, demand-side changes in food choices and consumption can help to achieve global GHG
32 mitigation targets (*high confidence*). Low-carbon diets on average tend to be healthier and have
33 smaller land footprints. By 2050, technical mitigation potential of dietary changes range from 2.7–6.4
34 GtCO₂-eq yr⁻¹ for assessed diets (*high confidence*). At the same time, the economic potential of such
35 solutions is lower, ranging from 1.8–3.4 Gt CO₂-eq yr⁻¹ at prices of 20–100 USD/tCO₂, with caloric
36 costs up to 190 kcal cap⁻¹ day⁻¹. The feasibility of how to create economically viable transitions to
37 more sustainable and healthy diets that also respect food security requirements needs to be addressed
38 in future research.

39

40 5.5.2.2 *Role of dietary preferences*

41 Food preference is an inherently cultural dimension that can ease or hinder transformations to food
42 systems that contribute to climate change mitigation. Consumer choice and dietary preferences are
43 guided by social, cultural, environmental, and traditional factors as well as economic growth. The
44 food consumed by a given group conveys cultural significance about social hierarchy, social systems
45 and human-environment relationships (Herforth and Ahmed 2015).

1 As suggested by Springmann et al. (2018a), per capita dietary emissions will translate into different
2 realised diets, according to regional contexts including cultural and gendered norms (e.g. among some
3 groups, eating meat is perceived as more masculine (Ruby and Heine 2011). In some cases, women
4 and men have different preferences in terms of food, with women reporting eating healthier food
5 (Imamura et al. 2015; Kiefer et al. 2005; Fagerli and Wandel 1999): these studies found that men tend
6 to eat more meat, while women eat more vegetables, fruits and dairy products (Kanter and Caballero
7 2012).

8 Food preferences can change over time, with the nutrition transition from traditional diets to high-
9 meat, high-sugar, high-saturated fat diets being a clear example of significant changes occurring in a
10 short period of time. Meat consumption per capita consistently responds to income with a saturating
11 trend at high income levels (Sans and Combris 2015; Vranken et al. 2014). Some emerging economies
12 have rapidly increased demand for beef, leading to pressure on natural resources (Bowles et al. 2019).
13 In another example, by reducing beef consumption between 2005 and 2014, Americans avoided
14 approximately 271 million metric tonnes of emissions (CO₂eq) (NRDC 2017). See Section **Error!**
15 **reference source not found.** for quantitative analysis. Attending farmers markets or buying directly
16 from local producers has been shown to change worldviews (Kerton and Sinclair 2010), and food
17 habits towards healthier diets (Pascucci et al. 2011) can be advanced through active learning (Milestad
18 et al. 2010).

19 Regarding the options to reduce meat intake in developed countries, research shows that there is an
20 apparent sympathy of consumers for meat reduction due to environmental impacts (Dagevos and
21 Voordouw 2013) which has not been exploited. Social factors that influence reducing meat
22 consumption in New Zealand include the need for better education or information dispersal regarding
23 perceived barriers to producing meat-reduced/less meals; ensuring there is sensory or aesthetic appeal;
24 and placing emphasis on human health or nutritional benefits (Tucker 2018).

25 Different and complementary strategies can be used in parallel for different consumer's profiles to
26 facilitate step-by-step changes in the amounts and the sources of protein consumed. In the
27 Netherlands, a nationwide sample of 1083 consumers were used to study their dietary choices toward
28 smaller portions of meat, smaller portions using meat raised in a more sustainable manner, smaller
29 portions and eating more vegetable protein, and meatless meals with or without meat substitutes.
30 Results showed that strategies to change meat eating frequencies and meat portion sizes appeared to
31 overlap and that these strategies can be applied to address consumers in terms of their own
32 preferences (de Boer et al. 2014).

34 **5.5.2.3 Uncertainties in demand-side mitigation potential**

35 Both reducing ruminant meat consumption and increasing its efficiency are often identified as main
36 options to reduce greenhouse gases emissions (GHGE) and to lessen pressure on land (Westhoek et al.
37 2014) (See Section 5.6 for synergies and trade-offs with health and Section 5.7 for discussion of Just
38 Transitions). However, analysing ruminant meat production is highly complex because of the extreme
39 heterogeneity of production systems and due to the numerous products and services associated with
40 ruminants (Gerber et al. 2015). See Supplementary Material Section SM5.5 for further discussion of
41 uncertainties in estimates of livestock mitigation technical potential. Further, current market
42 mechanisms are regarded as insufficient to decrease consumption or increase efficiency, and
43 governmental intervention is often suggested to encourage mitigation in both the supply-side and
44 demand-side of the food system (See Section 5.7) (Wirsenius et al. 2011; Henderson et al. 2018).

45 Minimising GHG emissions through mathematical programming with near-minimal acceptability
46 constraints can be understood as a reference or technical potential for mitigation through diet shifts. In
47 this context (Macdiarmid et al. 2012) found up to 36% reduction in emissions in UK with similar diet

1 costs applying fixed lifecycle analyses (LCA) carbon footprints (i.e., no rebound effects considered).
2 Westhoek et al. (2014) found 25-40 % in emissions by halving meat, dairy and eggs intake in EU,
3 applying standard IPCC fixed emission intensity factors. Uncertainty about the consequences of on-
4 the-ground implementation of policies towards low ruminant meat consumption in the food system
5 and their externalities remain noteworthy.

6 Often, all emissions are allocated to only to human edible meat and the boundaries are set only within
7 the farm gate (Henderson et al. 2018; Gerber et al. 2013). However, less than 50% of slaughtered
8 cattle weight is human edible meat, and 1-10% of the mass is lost or incinerated, depending on
9 specified risk materials legislation. The remaining mass provide inputs to multiple industries e.g.
10 clothing, furniture, vehicle coating materials, biofuel, gelatine, soap, cosmetics, chemical and
11 pharmaceutical industrial supplies, pet feed ingredients and fertilisers (Marti et al. 2011; Mogensen et
12 al. 2016; Sousa et al. 2017). This makes ruminant meat production one of the most complex problems
13 for LCA in the food system (Place and Mitloehner 2012; de Boer et al. 2011). There are only a few
14 examples taking into account slaughter byproducts e.g., Mogensen et al. (2016).

16 **5.5.2.4 Insect-based diets**

17 Edible insects are, in general, rich in protein, fat, and energy and can be a significant source of
18 vitamins and minerals (Rumpold and Schlüter 2015). Approximately 1,900 insect species are eaten
19 worldwide, mainly in developing countries (van Huis 2013). The development of safe rearing and
20 effective processing methods are mandatory for utilisation of insects in food and feed. Some insect
21 species can be grown on organic side streams, reducing environmental contamination and
22 transforming waste into high-protein feed. Insects are principally considered as meat substitutes, but
23 worldwide meat substitute consumption is still very low, principally due to differences in food
24 culture, and will require transition phases such as powdered forms (Megido et al. 2016; Smetana et al.
25 2015). Wider consumer acceptability will relate to pricing, perceived environmental benefits, and the
26 development of tasty insect-derived protein products (van Huis et al. 2015; van Huis 2013). Clearly
27 increasing share of insect-derived protein has the potential to reduce GHG emissions otherwise
28 associated with livestock production. No study to date however has quantified such potential.

30 **5.5.2.5 Food loss and waste, food security, and land use**

31 Food loss and waste impacts food security by reducing global and local food availability, limiting
32 food access due to increase in food price and decrease of producer income, and affecting future food
33 production due to unsustainable use of natural resources (HLPE 2014). Food loss is defined as the
34 reduction of edible food during production, postharvest, and processing, whereas food discarded by
35 consumers is considered as food waste (FAO 2011b). Combined food loss and waste amount to a third
36 of global food production (*high confidence*). During 2010-2016, global food loss and waste equalled
37 8–10% of total GHG emissions (*medium confidence*); and cost about USD 1 trillion per year (FAO
38 2014b) (*low confidence*).

39 A large share of produced food is lost in developing countries due to poor infrastructure, while a large
40 share of produced food is wasted in developed countries (Godfray et al. 2010). Changing consumer
41 behaviour to reduce per capita overconsumption offers substantial potential to improve food security
42 by avoiding related health burdens (Alexander et al. 2017; Smith 2013) and reduce emissions
43 associated with the extra food (Godfray et al. 2010). In 2007, around 20% of the food produced went
44 to waste in Europe and North America, while around 30% of the food produced was lost in sub-
45 Saharan Africa (FAO 2011b). During the last 50 years, the global food loss and waste increased from
46 around 540 Mt in 1961 to 1630 Mt in 2011 (Porter et al. 2016).

1 In 2011, food loss and waste resulted in about 8–10% of the total anthropogenic greenhouse gas
2 emissions of the entire food system. The mitigation potential of reduced food loss and waste from a
3 full life-cycle perspective, i.e., considering both food supply chain activities and land use change, was
4 estimated as 4.4 Gt CO₂-eq yr⁻¹ (FAO 2015a, 2013b). At a global scale, loss and waste of milk,
5 poultry meat, pig meat, sheep meat, and potatoes are associated with 3% of the global agricultural
6 N₂O emissions (more than 200 Gg N₂O-N yr⁻¹ or 0.06 Gt CO₂-eq yr⁻¹) in 2009 (Reay et al. 2012). For
7 the United States, 35% of energy use, 34% of blue water use, 34% of GHG emissions, 31% of land
8 use, and 35% of fertiliser use related to an individual's food-related resource consumption were
9 accounted for as food waste and loss in 2010 (Birney et al. 2017).

10 Similar to food waste, overconsumption, defined as food consumption in excess of nutrient
11 requirements, leads to GHG emissions (Alexander et al. 2017). In Australia for example,
12 overconsumption accounts for about 33% GHGs associated with food (Hadjikakou 2017). In addition
13 to GHG emissions, overconsumption also can lead to severe health conditions such as obesity or
14 diabetes. Over-eating was found to be at least as large a contributor to food system losses (Alexander
15 et al. 2017). Similarly, food system losses associated with consuming resource-intensive animal-based
16 products instead of nutritionally-comparable plant-based alternatives are defined as 'opportunity food
17 losses.' These were estimated to be 96, 90, 75, 50, and 40% for beef, pork, dairy, poultry, and eggs,
18 respectively, in the US (Shepon et al. 2018).

19 Avoiding food loss and waste will contribute to reducing emissions from the agriculture sector. By
20 2050, agricultural GHG emissions associated with production of food that might be wasted may
21 increase to 1.9–2.5 Gt CO₂-eq yr⁻¹ (Hiç et al. 2016). When land use change for agriculture expansion
22 is also considered, halving food loss and waste reduces the global need for cropland area by around
23 14% and GHG emissions from agriculture and land use change by 22–28% (4.5 Gt CO₂-eq yr⁻¹)
24 compared to the baseline scenarios by 2050 (Bajželj et al. 2014). The GHG emissions mitigation
25 potential of food loss and waste reduction would further increase when life cycle analysis accounts for
26 emissions throughout food loss and waste through all food system activities.

27 Reducing food loss and waste to zero might not be feasible. Therefore, appropriate options for the
28 prevention and management of food waste can be deployed to reduce food loss and waste and to
29 minimise its environmental consequences. Papargyropoulou et al. (2014) proposed the 3Rs (i.e.,
30 reduction, recovery and recycle) options to prevent and manage food loss and waste. A wide range of
31 approaches across the food supply chain is available to reduce food loss and waste, consisting of
32 technical and non-technical solutions (Lipinski et al. 2013). However, technical solutions (e.g.,
33 improved harvesting techniques, on-farm storage, infrastructure, packaging to keep food fresher for
34 longer, etc.) include additional costs (Rosegrant et al. 2015) and may have impacts on local
35 environments (FAO 2018b). Additionally, all parts of food supply chains need to become efficient to
36 achieve the full reduction potential of food loss and waste (Lipinski et al. 2013).

37 Together with technical solutions, approaches (i.e, non- technical solutions) to changes in behaviours
38 and attitudes of a wide range of stakeholders across the food system will play an important role in
39 reducing food loss and waste. Food loss and waste can be recovered by distributing food surplus to
40 groups affected by food poverty or converting food waste to animal feed (Vandermeersch et al. 2014).
41 Unavoidable food waste can also be recycled to produce energy based on biological, thermal and
42 thermochemical technologies (Pham et al. 2015). Additionally, strategies for reducing food loss and
43 waste also need to consider gender dynamics with participation of females throughout the food supply
44 chain (FAO 2018f).

45 In summary, reduction of food loss and waste can be considered as a climate change mitigation
46 measure that provides synergies with food security and land use (*robust evidence, medium*
47 *agreement*). Reducing food loss and waste reduces agricultural GHG emissions and the need for
48 agricultural expansion for producing excess food. Technical options for reduction of food loss and

1 waste include improved harvesting techniques, on-farm storage, infrastructure, and packaging.
2 However, the beneficial effects of reducing food loss and waste will vary between producers and
3 consumers, and across regions. Causes of food loss (e.g., lack of refrigeration) and waste (e.g.,
4 behaviour) differ substantially in developed and developing countries (*robust evidence, medium*
5 *agreement*). Additionally, food loss and waste cannot be avoided completely.

7 **5.5.2.6 Shortening supply chains**

8 Encouraging consumption of locally produced food and enhancing efficiency of food processing and
9 transportation can in some cases minimise food loss, contribute to food security, and reduce GHG
10 emissions associated with energy consumption and food loss. For example, Michalský and Hooda
11 (2015), through a quantitative assessment of GHG emissions of selected fruits and vegetables in the
12 UK, reported that increased local production offers considerable emissions savings. They also
13 highlighted that when imports are necessary, importing from Europe instead of the Global South can
14 contribute to considerable GHG emissions savings. Similar results were found by Audsley et al.
15 (2010), with exceptions for some foods, such as tomatoes, peppers or sheep and goat meat. Similarly,
16 a study in India shows that long and fragmented supply chains, which lead to disrupted price signals,
17 unequal power relations perverse incentives and long transport time, could be a key barrier to
18 reducing post-harvest losses (CIPHET 2007).

19 In other cases, environmental benefits associated with local food can be offset by inefficient
20 production systems with high emission intensity and resource needs, e.g., water, due to local
21 conditions. For example, vegetables produced in open fields can have much lower GHG emissions
22 than locally produced vegetables from heated greenhouses (Theurl et al. 2014). Whether locally
23 grown food has a lower carbon footprint depends on the on-farm emissions intensity as well as the
24 transport emissions. In some cases, imported food may have a lower carbon footprint than locally
25 grown food because some distant countries can produce food at much lower emissions intensity. For
26 example, Avetisyan et al. (2014) reported that regional variation of emission intensities associated
27 with production of ruminant products have large implications for emissions associated with local
28 food. They showed that consumption of local livestock products can reduce emissions due to short
29 supply chains in countries with low emission intensities; however, this might not be the case in
30 countries with high emission intensities.

31 In addition to improving emission intensity, efficient distribution systems for local food are needed
32 for lowering carbon footprints (Newman et al. 2013). Emissions associated with food transport
33 depend on the mode of transport, for example, emissions are lower for rail rather than truck (Brodt et
34 al. 2013). Tobarra et al. (2018) reported that emissions saving from local food may vary across
35 seasons and regions of import. They highlighted that in Spain local production of fruits and vegetables
36 can reduce emissions associated with imports from Africa but imports from France and Portugal can
37 save emissions in comparison to production in Spain. Additionally, local production of seasonal
38 products in Spain reduces emissions, while imports of out-of-season products can save emissions
39 rather than producing them locally.

40 In summary, consuming locally grown foods can reduce GHG emissions, if they are grown efficiently
41 (*high confidence*). The emissions reduction potential varies by region and season. Whether food with
42 shorter supply chains has a lower carbon footprint depends on both the on-farm emissions intensity as
43 well as the transport emissions. In some cases, imported food may have a lower carbon footprint
44 because some distant agricultural regions can produce food at lower emissions intensities.

1 **5.6 Mitigation, Adaptation, Food Security, and Land Use – Synergies,** 2 **Trade-Offs, and Co-Benefits**

3 Food systems will need to adapt to changing climates and also to reduce their greenhouse gas (GHG)
4 emissions and sequester carbon if Paris Agreement goals are to be met (Springmann et al. 2018a; van
5 Vuuren et al. 2014). The synergies and trade-offs between the food system mitigation and adaptation
6 options described in Sections 5.3 and 5.5 are of increasing importance in both scientific and policy
7 communities because of the necessity to ensure food security, i.e., provision of nutritious food for the
8 growing population, while responding to climate change (Rosenzweig and Hillel 2015) A special
9 challenge involves interactions between land-based non-food system mitigation, such as negative
10 emissions technologies, and food security. Response options for the food system have synergies and
11 trade-offs between climate change mitigation and adaptation (Figure 5.13; Chapter 6).

12 Tirado et al. (2013) suggest an integrated approach to address the impacts of climate change to food
13 security that considers a combination of nutrition-sensitive adaptation and mitigation measures,
14 climate-resilient and nutrition-sensitive agricultural development, social protection, improved
15 maternal and child care and health, nutrition-sensitive risk reduction and management, community
16 development measures, nutrition-smart investments, increased policy coherence, and institutional and
17 cross-sectoral collaboration. These measures are a meansto achieve both short-term and long-term
18 benefits in poor and marginalised groups.

19 This section assesses the synergies and trade-offs for land-based atmospheric carbon dioxide removal
20 measures, effects of mitigation measures on food prices, and links between dietary choices and human
21 health. It then evaluates a range of integrated agricultural systems and practices that combine
22 mitigation and adaptation measures, including the role of agricultural intensification. The roles of
23 women’s empowerment and urban agriculture are examined, as well as interactions between SDG2
24 (Zero Hunger) and SDG 13 (Climate Action).

25

Food system response options

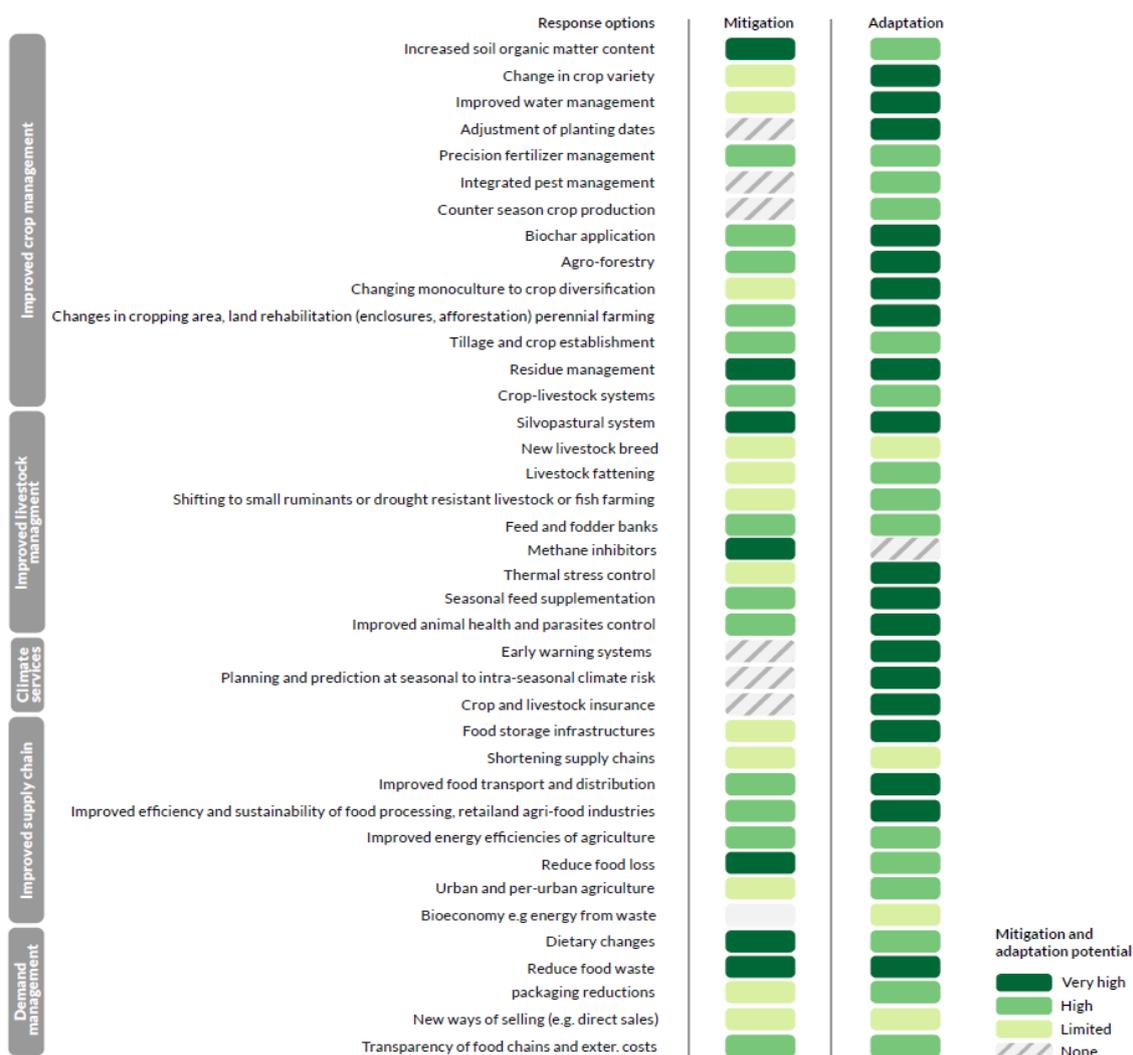


Figure 5.13 Response options related to food system and their potential impacts on mitigation and adaptation. Many response options offer significant potential for both mitigation and adaptation.

5.6.1 Land-based carbon dioxide removal (CDR) and bioenergy

Large-scale deployment of negative emission technologies (NETs) in emission scenarios has been identified as necessary for avoiding unacceptable climate change (IPCC 2018b). Among the available NETs, carbon dioxide removal (CDR) technologies are receiving increasing attention. Land-based CDRs include afforestation and reforestation (AR), sustainable forest management, biomass energy with carbon capture and storage (BECCS), and biochar (BC) production (Minx et al. 2018). Most of the literature on global land-based mitigation potential relies on CDRs, particularly on BECCS, as a major mitigation action (Kraxner et al. 2014; Larkin et al. 2018; Rogelj et al. 2018, 2015, 2011). BECCS is not yet deployable at a commercial scale, as it faces challenges similar to fossil fuel carbon capture and storage (CCS) (Fuss et al. 2016; Vaughan and Gough 2016; Nemet et al. 2018). Regardless, the effectiveness of large-scale BECCS to meet Paris Agreement goals has been questioned and other pathways to mitigation have been proposed (Anderson and Peters 2016; van Vuuren et al. 2017, 2018; Grubler et al. 2018; Vaughan and Gough 2016).

Atmospheric CO₂ removal by storage in vegetation depends on achieving net organic carbon accumulation in plant biomass over decadal time scales (Kemper 2015) and, after plant tissue decay, in soil organic matter (Del Grosso et al. 2019). AR, BECCS and BC differ in the use and storage of

1 plant biomass. In BECCS, biomass carbon from plants is used in industrial processes (e.g., for
2 electricity, hydrogen, ethanol, and biogas generation), releasing CO₂, which is then captured and
3 geologically stored (Greenberg et al. 2017; Minx et al. 2018).

4 Afforestation and reforestation result in long-term carbon storage in above and belowground plant
5 biomass on previously unforested areas, and is effective as a carbon sink during the AR
6 establishment period, in contrast to thousands of years for geological C storage (Smith et al. 2016).

7 Biochar is produced from controlled thermal decomposition of biomass in absence of oxygen
8 (pyrolysis), a process that also yields combustible oil and combustible gas in different proportions.
9 Biochar is a very stable carbon form, with storage on centennial timescales (Lehmann et al. 2006)
10 (See also Chapter 4). Incorporated in soils, some authors suggest it may lead to improved water-
11 holding capacity, nutrient retention, and microbial processes (Lehmann et al. 2015). There is
12 however considerable uncertainty about the benefits and risks of this practice (The Royal Society
13 2018).

14 Land-based CDRs require high biomass-producing crops. Since not all plant biomass is harvested
15 (e.g., roots and harvesting losses), it can produce co-benefits related to soil carbon sequestration, crop
16 productivity, crop quality, as well improvements in air quality, but the overall benefits strongly
17 depend on the previous land use and soil management practices (Smith et al. 2016; Wood et al. 2018).
18 In addition, CDR effectiveness varies widely depending on type of biomass, crop productivity, and
19 emissions offset in the energy system. Importantly, its mitigation benefits can be easily lost due to
20 land-use change interactions (Harper et al. 2018; Fuss et al. 2018; Daioglou et al. 2019).

21 Major common challenges of implementing these large-scale CDR solutions, as needed to stabilise
22 global temperature “well-below” 2°C by the end of the century, are the large investments and the
23 associated significant changes in land use required. Most of the existing scenarios estimate the
24 global area required for BECCS alone in the range of 109-990 Mha (IPCC 2018a), most commonly
25 around 380–700 Mha (Smith et al. 2016), reaching rates of net area expansion rates up to 23.7 M
26 ha yr⁻¹ (IPCC 2018b). The upper limit implies unprecedented rates of area expansion for crops and
27 forestry observed historically, for instance as reported by FAO since 1961 (FAOSTAT 2018). By
28 comparison, the sum of recent worldwide rates of expansion in harvested area of soybean and
29 sugarcane has not exceeded 3.5 M ha yr⁻¹ on average. Even at this rate, they have been the source
30 of major concerns for their possible negative environmental and food security impacts (Boerema et
31 al. 2016; Popp et al. 2014).

32 Most land area available for CDR is currently pasture, estimated at 3,300 Mha globally (FAOSTAT
33 2018). However, there is *low confidence* about how much low-productivity land is actually available
34 for CDR (Lambin et al. 2013; Gibbs and Salmon 2015). There is also *low confidence* and *low*
35 *agreement* if the transition to BECCS will take place directly on low-productivity grasslands
36 (Johansson and Azar 2007), and uncertainty on the governance mechanisms required to avoid
37 unwanted spill-over effects, for instance causing additional deforestation (Keles et al. 2018).

38 Further, grasslands and rangelands may often occur in marginal areas, in which case they may be
39 exposed to climate risks, including periodic flooding. Grasslands and especially rangelands and
40 savannahs tend to predominate in less-developed regions, often bordering areas of natural vegetation
41 with little infrastructure available for transport and processing of large quantities of CDR-generated
42 biomass (O’Mara 2012; Beringer et al. 2011; Haberl et al. 2010; Magdoff 2007).

43 CDR-driven reductions in available pastureland area is a scenario of constant or increasing global
44 animal protein output as proposed by (Searchinger et al. 2018). However, despite the recent reduction
45 in meat consumption in western countries, this will require productivity improvements (Cohn et al.
46 2014; Strassburg et al. 2014). It would also result in lower emission intensities and create conditions
47 for increased soil carbon stocks (de Oliveira Silva et al. 2016; Searchinger et al. 2018; Soussana et al.

1 2019, 2013). At the same time, food security may be threatened if land-based mitigation displaced
2 crops elsewhere, especially if to regions of lower productivity potential, higher climatic risk, and
3 higher vulnerability.

4 There is *low agreement* about what are the more competitive regions of the world for CDRs. Smith et
5 al. (2016) and Vaughan et al. (2018) identify as candidates relatively poor countries in Latin America,
6 Africa and Asia (except China and India). Others indicate those regions may be more competitive for
7 food production, placing Europe as a major BECCS exporter (Muratori et al. 2016). Economically
8 feasible CDR investments are forecast to be directed to regions with high biomass production
9 potential, demand for extra energy production, low leakage potential for deforestation and low
10 competition for food production (Vaughan et al. 2018). Latin America and Africa, for instance,
11 although having high biomass production potential, still have low domestic energy consumption (589
12 and 673 MTOE – 24.7 and 28.2 EJ, respectively), with about 30% of primary energy from renewable
13 sources (reaching 50% in Brazil), mainly hydropower and traditional biomass.

14 There is *high confidence* that deployment of BECCS will require ambitious investments and policy
15 interventions (Peters and Geden 2017) with strong regulation and governance of bioenergy production
16 to ensure protection of forests, maintain food security and enhance climate benefits (Burns and
17 Nicholson 2017; Vaughan et al. 2018; Muratori et al. 2016), and that such conditions may be
18 challenging for developing countries. Increased value of bioenergy puts pressure on land, ecosystem
19 services, and the prices of agricultural commodities, including food (*high confidence*).

20 There is *medium confidence* for the impact of CDR technologies on increased food prices and reduced
21 food security, as these depend on several assumptions. Nevertheless, those impacts could be strong,
22 with food prices doubling under certain scenario combinations (Popp et al. 2017). The impacts of
23 land-mitigation policies on the reduction of dietary energy availability alone, i.e., without climate
24 change impacts, is estimated at over 100 kcal.person⁻¹ day⁻¹ by 2050, with highest regional impacts in
25 south Asia and sub-Saharan Africa (Hasegawa et al. 2018) (See Section 5.2). However, only limited
26 pilot BECCS projects have been implemented to date (Lenzi et al. 2018). Integrated assessment
27 models (IAMs) use theoretical data based on high-level studies and limited regional data from the few
28 on-the-ground BECCS projects.

29 Furthermore, it has been suggested that several BECCS IAM scenarios rely on unrealistic
30 assumptions regarding regional climate, soils and infrastructure suitability (Anderson and Peters
31 2016), as well as international bioenergy trade (Lamers et al. 2011). Current global IAMs usually
32 consider major trends in production potential and projected demand, overlooking major challenges for
33 the development of a reliable international market. Such a market will have to be created from scratch
34 and overcome a series of constraints, including trade barriers, logistics, and supply chains, as well as
35 social, ecological and economic impacts (Matzenberger et al. 2015).

36 In summary, there is *high agreement* that better assessment of BECCS mitigation potential would
37 need to be based on increased regional, bottom-up studies of biomass potentials, socio-economic
38 consequences (including on food security), and environmental impacts in order to develop more
39 realistic estimates (IPCC 2018a).

40

41 **5.6.2 Mitigation, food prices, and food security**

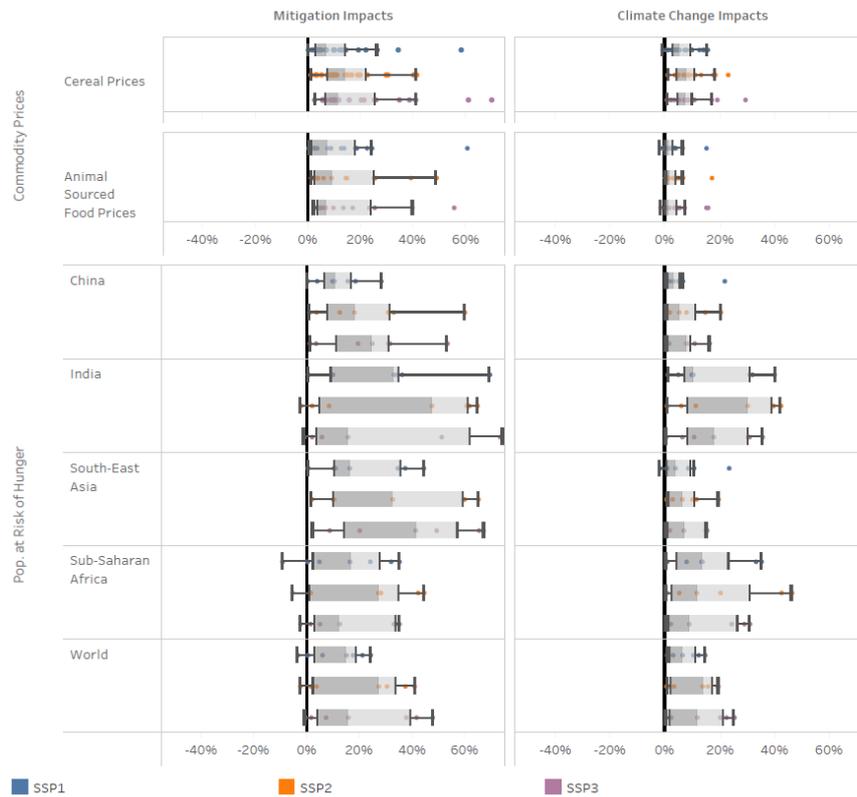
42 Food prices are the result of supply, demand and trade relations. Earlier studies (e.g., (Nelson et al.
43 2009)) showed that recent climate impacts that reduced crop productivity led to higher prices and
44 increased trade of commodities between regions, with asymmetric impacts on producers and
45 consumers. In terms of published scenario analyses, the most affected regions tend to be Sub-Saharan
46 Africa and parts of Asia, but there is significant heterogeneity in results between countries. Relocation

1 of production to less affected areas buffers these impacts to a certain extent, as well as potential for
2 improvements in food production technologies (Hasegawa et al. 2018; van Meijl et al. 2017; Wiebe et
3 al. 2015; Lotze-Campen et al. 2014; Valin et al. 2014; Robinson et al. 2014).

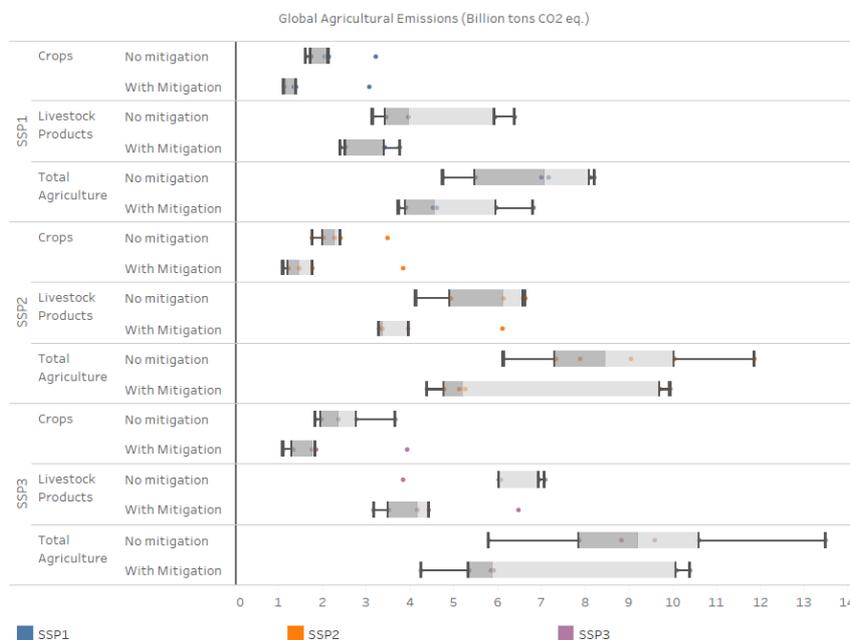
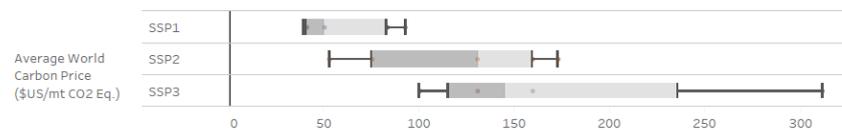
4 A newer, less studied impact of climate change on prices and their impacts on food security is the
5 level of land-based mitigation necessary to stabilise global temperature. Hasegawa et al. (2018), using
6 an ensemble of seven global economic models across a range of greenhouse gas emissions pathways
7 and socioeconomic trajectories, suggested that the level of mitigation effort needed to reduce
8 emissions can have a more significant impact on prices than the climate impacts themselves on
9 reduced crop yields (Figure 5.14). This occurs because in the models, taxing GHG emissions leads to
10 higher crop and livestock prices, while land-based mitigation leads to less land availability for food
11 production, potentially lower food supply, and therefore food price increases.

12 Price increases in turn lead to reduced consumption, especially by vulnerable groups, or to shifts
13 towards cheaper food, which are often less nutritious. This leads to significant increases in the number
14 of malnourished people. Frank et al. (2017) and Fujimori et al. (2017) arrived at the same conclusions
15 for the 1.5°C mitigation scenario using the IAM Globiom and ensembles of AgMIP global economic
16 models. While the magnitude of the response differs between models, the results are consistent
17 between them. In contrast, a study based on five global agro-economic models highlights that the
18 global food prices may not increase much when the required land for bioenergy is accessible on the
19 margin of current cropland, or the feedstock does not have a direct competition with agricultural land
20 (Lotze-Campen et al. 2014).

21 These studies highlight the need for careful design of emissions mitigation policies in upcoming
22 decades—for example, targeted schemes encouraging more productive and resilient agricultural
23 production systems and the importance of incorporating complementary policies (such as safety-net
24 programmes for poverty alleviation) that compensate or counteract the impacts of the climate change
25 mitigation policies on vulnerable regions (Hasegawa et al. 2018). Fujimori et al. (2018) showed how
26 an inclusive policy design can avoid adverse side-effects on food security through international aid,
27 bioenergy taxes, or domestic reallocation of income. These strategies can shield impoverished and
28 vulnerable people from the additional risk of hunger that would be caused by the economic effects of
29 policies narrowly focussing on climate objectives only.



1



2

3 **Figure 5.14 Regional impacts of climate change and mitigation on food price (top), population (pop) at**
 4 **risk of hunger or undernourishment (middle), greenhouse gas emissions (bottom) in 2050 under different**
 5 **socio-economic scenarios (SSP1, SSP2 and SSP3). Values indicate changes from no climate change and**
 6 **no climate change mitigation scenario. MAGPIE, a global land use allocation model, is excluded due to**

1 **inelastic food demand. The value of India includes that of Other Asia in MAGNET, a global general**
2 **equilibrium model (Hasegawa et al. 2018)**

3 In summary, food security will be threatened through increasing numbers of malnourished people if
4 land-based mitigation raises prices, unless other policy mechanisms reduce its impact (*high*
5 *confidence*). Inclusive policy design can avoid adverse side-effects on food security by shielding
6 vulnerable people from the additional risk of hunger that would be caused by the economic effects of
7 policies narrowly focusing on climate objectives (*medium confidence*).

8

9 **5.6.3 Environmental and health effects of adopting healthy and sustainable diets**

10 Two key questions arise from the potentially significant mitigation potential of dietary change: 1) Are
11 ‘low-GHG emission diets’ likely to be beneficial for health? and 2) Would changing diets at scale
12 provide substantial benefits? In short, what are the likely synergies and trade-offs between low-GHG
13 emissions diets and food security, health, and climate change? See Supplementary Material Section
14 SM5.6 for further discussion.

15 *Are “low GHG emission diets” healthy?* Consistent evidence indicates that, in general, a dietary
16 pattern that is higher in plant-based foods, such as vegetables, fruits, whole grains, legumes, nuts, and
17 seeds, and lower in animal-based foods, is more health-promoting and associated with lower
18 environmental impact (GHG emissions and energy, land, and water use) than either the current global
19 average diets (Swinburn et al. 2019; Willett et al. 2019; Springmann et al. 2016b), or the current
20 average US diet (Nelson et al. 2016). Another study (Van Mierlo et al. 2017) showed that
21 nutritionally-equivalent diets can substitute plant-based foods for meat and provide reductions in
22 GHG emissions.

23 There are several studies that estimate health adequacy and sustainability and conclude that healthy
24 sustainable diets are possible. These include global studies (e.g., (Willett et al. 2019; Swinburn et al.
25 2019)), as well as localised studies (e.g., (Van Dooren et al. 2014). For example, halving
26 consumption of meat, dairy products and eggs in the European Union would achieve a 40% reduction
27 in ammonia emissions, 25–40% reduction in non-CO₂ GHG emissions (primarily from agriculture)
28 and 23% per capita less use of cropland for food production, with dietary changes lowering health
29 risks (Westhoek et al. 2014). In China, diets were designed that could meet dietary guidelines while
30 creating significant reductions in GHG emissions (between 5% and 28%, depending on scenario)
31 (Song et al. 2017). Changing diets can also reduce non-dietary related health issues caused by
32 emissions of air pollutants; for example, specific changes in diets were assessed for their potential to
33 mitigate PM_{2.5} in China (Zhao et al. 2017b).

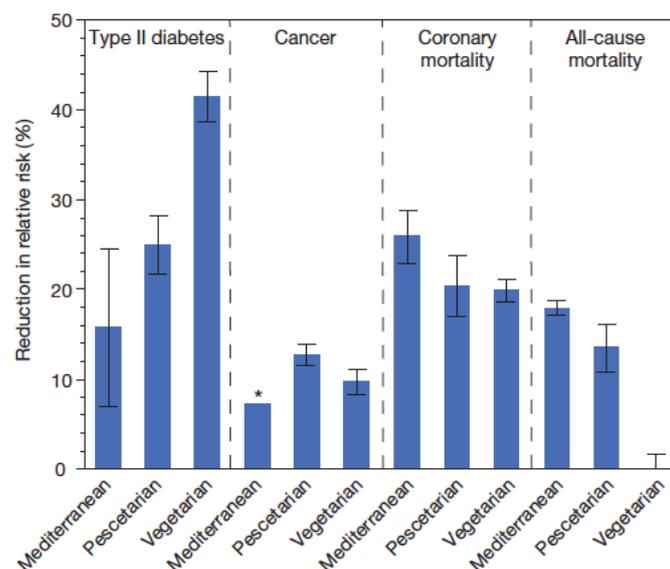
34 A range of studies are starting to estimate both health and environmental benefits from dietary shifts.
35 For example, (Farchi et al. 2017) estimate health (colorectal cancer, cardio-vascular disease) and
36 GHG reductions of “Mediterranean” diets, low in meat, in Italy, finding the potential to reduce deaths
37 from colorectal cancer of 7-10% and CVD from 9-10%, and potential savings of up to 263 CO₂-
38 eq.person⁻¹.yr⁻¹. In the US, Hallström et al. (2017) found that adoption of healthier diets (consistent
39 with dietary guidelines, and reducing amounts of red and processed meats) could reduce relative risk
40 of coronary heart disease, colorectal cancer, and type 2 diabetes by 20–45%, US health care costs by
41 USD 77–93 billion per year, and direct GHG emissions by 222–826 kg CO₂-eq/capita yr⁻¹ (69–84 kg
42 from the health care system, 153–742 kg from the food system). Broadly similar conclusions were
43 found for the Netherlands (Biesbroek et al. 2014); and the UK (Friel et al. 2009; Milner et al. 2015).

44 Whilst for any given disease, there are a range of factors, including diet, that can affect it, and
45 evidence is stronger for some diseases than others, a recent review found that an overall trend toward
46 increased cancer risk was associated with unhealthy dietary patterns, suggesting that diet-related

1 choices could significantly affect the risk of cancer (Grosso et al. 2017). Tilman and Clark (2014)
 2 found significant benefits in terms of reductions in relative risk of key diseases: type 2 diabetes,
 3 cancer, coronary mortality and all causes of mortality (Figure 5.11).

5.6.3.1 Can dietary shifts provide significant benefits?

6 Many studies now indicate that dietary shifts can significantly reduce GHG emissions. For instance,
 7 several studies highlight that if current dietary trends are maintained, this could lead to emissions
 8 from agriculture of approximately 20 Gt CO₂-eq yr⁻¹ by 2050, creating significant mitigation potential
 9 (Pradhan et al. 2013b; Bajželj et al. 2014; Hedenus et al. 2014; Bryngelsson et al. 2017). Additionally
 10 in the US, a shift in consumption towards a broadly healthier diet, combined with meeting the USDA
 11 and Environmental Protection Agency's 2030 food loss and waste reduction goals, could increase *per*
 12 *capita* food-related energy use by 12%, decrease blue water consumption by 4%, decrease green
 13 water use by 23%, decrease GHG emissions from food production by 11%, decrease GHG emissions
 14 from landfills by 20%, decrease land use by 32%, and increase fertiliser use by 12% (Birney et al.
 15 2017). This study, however, does not account for all potential routes to emissions, ignoring, for
 16 example, fertiliser use in feed production. Similar studies have been conducted, for China (Li et al.
 17 2016), where adoption of healthier diets and technology improvements have the potential to reduce
 18 food systems GHG emissions by >40% relative to those in 2010; and India (Green et al. 2017; Vetter
 19 et al. 2017), where alternative diet scenarios can affect emissions from the food system by -20 to
 20 +15%.



22
 23 **Figure 5.11 Diet and health effects of different consumption scenarios (Tilman and Clark 2014) (* reflects**
 24 **data from a single study, hence no error bars)**

25 Springmann et al (2018a) modelled the role of technology, waste reduction and dietary change in
 26 living within planetary boundaries (Rockström et al. 2009), with the climate change boundary being a
 27 66% chance of limiting warming to less than 2°C. They found that all are necessary for achievement
 28 of a sustainable food system. Their principal conclusion is that only by adopting a “flexitarian diet”,
 29 as a global average, would climate change be limited to under two degrees. Their definition of a
 30 flexitarian diet is fruits and vegetables, plant-based proteins, modest amounts of animal-based
 31 proteins, and limited amounts of red meat, refined sugar, saturated fats, and starchy foods.

1 Healthy and sustainable diets address both health and environmental concerns (Springmann et al.
2 2018b). There is high agreement that there are significant opportunities to achieve both objectives
3 simultaneously. Contrasting results of marginal GHG emissions, i.e. variations in emissions as a result
4 of variation in one or more dietary components, are found when comparing low to high emissions
5 self-selected diets (diets freely chosen by consumers). (Vieux et al. 2013) found self-selected
6 healthier diets with higher amounts of plant-based food products did not result in lower emissions,
7 while (Rose et al. 2019) found that the lowest emission diets analysed were lower in meat but
8 higher in oil, refined grains and added sugar. (Vieux et al. 2018) concluded that setting nutritional
9 goals with no consideration for the environment may increase GHG emissions (GHGE).

10 Tukker et al. (2011) also found a slight increase in emissions by shifting diets towards the European
11 dietary guidelines, even with lower meat consumption. Heller and Keoleian (2015) found a 12%
12 increase in GHGE when shifting to iso-caloric diets, i.e., diets with the same caloric intake of diets
13 currently consumed, following the US guidelines and a 1% decrease in GHGE when adjusting caloric
14 intake to recommended levels for moderate activity. There is scarce information on the marginal
15 GHGE that would be associated with following dietary guidelines in developing countries.

16 Some studies have found a modest mitigation potential of diet shifts when economic and biophysical
17 systems effects are taken into account in association with current dietary guidelines. Tukker et al.
18 (2011), considering economic rebound effects of diet shifts (i.e., part of the gains would be lost due to
19 increased use at lower prices), found maximum changes in emissions of the EU food system of 8%
20 (less than 2% of total EU emissions) when reducing meat consumption by 40 to 58%. Using an
21 economic optimisation model for studying carbon taxation in food but with adjustments of
22 agricultural production systems and commodity markets in Europe (Zech and Schneider 2019) found
23 a reduction of 0.41% in GHG emissions at a tax level of 50 USDt⁻¹CO₂eq. They estimate a leakage of
24 43% of the greenhouse gas emissions reduced by domestic consumption, (i.e., although reducing
25 emissions due to reducing consumption, around 43% of the emissions would not be reduced because
26 part of the production would be directed to exports).

27 Studying optimised beef production systems intensification technologies in a scenario of no
28 grasslands area expansion (de Oliveira Silva et al. 2016) found marginal GHG emissions to be
29 negligible in response to beef demand in the Brazilian Cerrado. This was because reducing
30 productivity would lead to increased emission intensities, cancelling out the effect of reduced
31 consumption.

32 In summary, there is significant potential mitigation (*high confidence*) arising from the adoption of
33 diets in line with dietary recommendations made on the basis of health. These are broadly similar
34 across most countries. These are typically capped by at the number of calories and higher in plant-
35 based foods, such as vegetables, fruits, whole grains, legumes, nuts and seeds, and lower in animal-
36 sourced foods, fats and sugar. Such diets have the potential to be both more sustainable and healthy
37 than alternative diets (but healthy diets are not necessarily sustainable and vice versa). The extent to
38 which the mitigation potential of dietary choices can be realised requires both climate change and
39 health being considered together. Socio-economic (prices, rebound effects), political, and cultural
40 contexts would require significant consideration to enable this mitigation potential to be realised.

41

42 **5.6.4 Sustainable integrated agricultural systems**

43 A range of integrated agricultural systems are being tested to evaluate synergies between mitigation
44 and adaptation and lead to low-carbon and climate-resilient pathways for sustainable food security
45 and ecosystem health (*robust evidence, medium agreement*). Integration refers to the use of practices
46 that enhance an agroecosystem's mitigation, resilience, and sustainability functions. These systems
47 follow holistic approaches with the objective of achieving biophysical, socio-cultural, and economic

1 benefits from land management systems (Sanz et al. 2017). These integrated systems may include
2 agroecology (FAO et al. 2018; Altieri et al. 2015), climate smart agriculture (FAO 2011c; Lipper et
3 al. 2014; Aggarwal et al. 2018), conservation agriculture (Aryal et al. 2016; Sapkota et al. 2015), and
4 sustainable intensification (FAO 2011d; Godfray 2015), among others.

5 Many of these systems are complementary in some of their practices, although they tend to be based
6 on different narratives (Wezel et al. 2015; Lampkin et al. 2015; Pimbert 2015). They have been tested
7 in various production systems around the world (Dinesh et al. 2017; Jat et al. 2016; Sapkota et al.
8 2015; Neufeldt et al. 2013). Many technical innovations, e.g., precision nutrient management
9 (Sapkota et al. 2014) and precision water management (Jat et al. 2015), can lead to both adaptation
10 and mitigation outcomes and even synergies; although negative adaptation and mitigation outcomes
11 (i.e., trade-offs) are often overlooked. Adaptation potential of ecologically intensive systems includes
12 crop diversification, maintaining local genetic diversity, animal integration, soil organic management,
13 water conservation and harvesting the role of microbial assemblages (See Section 5.3). Technical
14 innovations may encompass not only inputs reduction, but complete redesign of agricultural systems
15 (Altieri et al. 2017) and how knowledge is generated (Levidow et al. 2014), including social and
16 political transformations.

18 **5.6.4.1 Agroecology**

19 Agroecology (see Glossary) (Francis et al. 2003; Gliessman and Engles 2014; Gliessman 2018),
20 provides knowledge for their design and management, including social, economic, political, and
21 cultural dimensions (Dumont et al. 2016). It started with a focus at the farm level but has expanded to
22 include the range of food system activities (Benkeblia 2018). Agroecology builds systems resilience
23 through knowledge-intensive practices relying on traditional farming systems and co-generation of
24 new insights and information with stakeholders through participatory action research (Menéndez et al.
25 2013). It provides a multidimensional view of food systems within ecosystems, building on
26 indigenous and local knowledge (ILK) and co-evolving with the experiences of local people, available
27 natural resources, access to these resources, and ability to share and pass on knowledge among
28 communities and generations, emphasising the inter-relatedness of all agroecosystem components and
29 the complex dynamics of ecological processes (Vandermeer 1995).

30 At the farm level, agroecological practices recycle biomass and regenerate soil biotic activities. They
31 strive to attain balance in nutrient flows to secure favorable soil and plant growth conditions,
32 minimise loss of water and nutrients, and improve use of solar radiation. Practices include efficient
33 microclimate management, soil cover, appropriate planting time and genetic diversity. They seek to
34 promote ecological processes and services such as nutrient cycling, balanced predator/prey
35 interactions, competition, symbiosis, and successional changes. The overall goal is to benefit human
36 and non-human communities in the ecological sphere, with fewer negative environmental or social
37 impacts and fewer external inputs (Vandermeer et al. 1998; Altieri et al. 1998). From a food system
38 focus, agroecology provides management options in terms of commercialisation and consumption
39 through the promotion of short food chains and healthy diets (Pimbert and Lemke 2018; Loconto et al.
40 2018).

41 Agroecology has been proposed as a key set of practices in building climate resilience (FAO et al.
42 2018; Altieri et al. 2015). These can enhance on-farm diversity (of genes, species, and ecosystems)
43 through a landscape approach (FAO 2018g). Outcomes include soil conservation and restoration and
44 thus soil carbon sequestration, reduction of the use of mineral and chemical fertilisers, watershed
45 protection, promotion of local food systems, waste reduction, and fair access to healthy food through
46 nutritious and diversified diets (Pimbert and Lemke 2018; Kremen et al. 2012; Goh 2011; Gliessman
47 and Engles 2014).

1 A principle agroecology is to contribute to food production by smallholder farmers (Altieri 2002).
2 Since climatic events can severely impact smallholder farmers, there is a need to better understand the
3 heterogeneity of small-scale agriculture in order to consider the diversity of strategies that traditional
4 farmers have used and still use to deal with climatic variability. In Africa, many smallholder farmers
5 cope with and even prepare for climate extremes, minimising crop failure through a series of
6 agroecological practices (e.g., biodiversification, soil management, and water harvesting) (Mbow et
7 al. 2014a). Resilience to extreme climate events is also linked to on-farm biodiversity, a typical
8 feature of traditional farming systems (Altieri and Nicholls 2017).

9 Critiques of agroecology refer to its explicit exclusion of modern biotechnology (Kershen 2013) and
10 the assumption that smallholder farmers are a uniform unit with no heterogeneity in power (and thus
11 gender) relationships (Neira and Montiel 2013; Siliprandi and Zuluaga Sánchez 2014).

13 **5.6.4.2 Climate-smart agriculture**

14 ‘Climate-smart agriculture’ (CSA) is an approach developed to tackle current food security and
15 climate change challenges in a joint and synergistic fashion (Lipper et al. 2014; Aggarwal et al. 2018;
16 FAO 2013c). CSA is designed to be a pathway towards development and food security built on three
17 pillars: increasing productivity and incomes, enhancing resilience of livelihoods and ecosystems and
18 reducing, and removing GHG emissions from the atmosphere (FAO 2013c). Climate-smart
19 agricultural systems are integrated approaches to the closely linked challenges of food security,
20 development, and climate change adaptation/mitigation to enable countries to identify options with
21 maximum benefits and those where trade-offs need management.

22 Many agricultural practices and technologies already provide proven benefits to farmers’ food
23 security, resilience and productivity (Dhanush and Vermeulen 2016). In many cases these can be
24 made implemented by changing the suites of management practices. For example, enhancing soil
25 organic matter to improve water-holding capacity of agricultural landscapes also sequesters carbon. In
26 annual cropping systems, changes from conventional tillage practices to minimum tillage can convert
27 the system from one that either provides only adaptation or mitigation benefits or neither types of
28 benefits to one that provides both adaptation and mitigation benefits (Sapkota et al. 2017a; Harvey et
29 al. 2014a).

30 Increasing food production by using more fertilisers in agricultural fields could maintain crop yield in
31 the face of climate change, but may result in greater overall GHG emissions. But increasing or
32 maintaining the same level of yield by increasing nutrient-use-efficiency through adoption of better
33 fertiliser management practices could contribute to both food security and climate change mitigation
34 (Sapkota et al. 2017a).

35 Mixed farming systems integrating crops, livestock, fisheries and agro-forestry could maintain crop
36 yield in the face of climate change, help the system to adapt to climatic risk, and minimise GHG
37 emissions by increasingly improving the nutrient flow in the system (Mbow et al. 2014a; Newaj et al.
38 2016; Bioversity International 2016). Such systems can help diversify production and/or incomes and
39 support efficient and timely use of inputs thus contributing to increased resilience, but require local
40 seed and input systems and extension services. Recent whole farm modelling exercises have shown
41 the economic and environmental (reduced GH emissions, reduced land use) benefits of integrated
42 crop-livestock systems. Gil et al. (2018) compared different soy-livestock systems across multiple
43 economic and environmental indicators, including climate resilience. However it is important to note
44 that potential benefits are very context specific.

45 Although climate-smart agriculture involves a holistic approach, some argue that it narrowly focuses
46 on technical aspects at the production level (Taylor 2018; Newell and Taylor 2018). Studying barriers
47 to the adoption and diffusion of technological innovations for climate-smart agriculture in Europe,

1 Long et al. (2016) found that there was incompatibility between existing policies and climate-smart
2 agriculture objectives, including barriers to the adoption of technological innovations.

3 Climate-smart agricultural systems recognise that the implementation of the potential options will be
4 shaped by specific country contexts and capacities, as well as enabled by access to better information,
5 aligned policies, coordinated institutional arrangements and flexible incentives and financing
6 mechanisms (Aggarwal et al. 2018). Attention to underlying socio-economic factors that affect
7 adoption of practices and access to technologies is crucial for enhancing biophysical processes,
8 increasing productivity, and reducing GHG emissions at scale. The Government of India, for example,
9 has started a program of climate resilient villages (CRV) as a learning platform to design, implement,
10 evaluate and promote various climate-smart agricultural interventions, with the goal of ensuring
11 enabling mechanisms at the community level (Srinivasa Rao et al. 2016).

12

13 **5.6.4.3 Conservation agriculture**

14 Conservation agriculture (CA) is based on the principles of minimum soil disturbance and permanent
15 soil cover combined with appropriate crop rotation (Jat et al. 2014; FAO 2011e). CA has been shown
16 to respond with positive benefits to smallholder farmers under both economic and environmental
17 pressures (Sapkota et al. 2017a, 2015). This agricultural production system uses a body of soil and
18 residues management practices that control erosion (Blanco Sepúlveda and Aguilar Carrillo 2016)
19 and at the same time to improve soil quality, by increasing organic matter content and improving
20 porosity, structural stability, infiltration and water retention (Sapkota et al. 2017a, 2015; Govaerts et
21 al. 2009)

22 Intensive agriculture during the second half of the 20th century led to soil degradation and loss of
23 natural resources and contributed to climate change. Sustainable soil management practices can
24 address both food security and climate change challenges faced by these agricultural systems. For
25 example, sequestration of soil organic carbon (SOC) is an important strategy to improve soil quality
26 and to mitigation of climate change (Lal 2004). CA has been reported to increase farm productivity by
27 reducing costs of production (Aryal et al. 2015; Sapkota et al. 2015; Indoria et al. 2017) as well as to
28 reduce GHG emission (Pratibha et al. 2016).

29 CA brings favourable changes in soil properties that affect the delivery of nature's contribution to
30 people (NCPs) or ecosystem services including climate regulation through carbon sequestration and
31 GHG emissions (Palm et al. 2013; Sapkota et al. 2017a). However, by analysing datasets for soil
32 carbon in the tropics, Powlson et al. (2014, 2016) argued that the rate of SOC increase and resulting
33 GHG mitigation in CA systems, from zero-tillage in particular, has been overstated (see also Chapter
34 2).

35 However, there is unanimous agreement that the gain in SOC and its contribution to GHG mitigation
36 by CA in any given soil is largely determined by the quantity of organic matter returned to the soil
37 (Giller et al. 2009; Virto et al. 2011; Sapkota et al. 2017b). Thus, a careful analysis of the production
38 system is necessary to minimise the trade-offs among the multiple use of residues, especially where
39 residues remain an integral part of livestock feeding (Sapkota et al. 2017b). Similarly, replacing
40 mono-cropping systems with more diversified cropping systems and agroforestry as well as
41 afforestation and deforestation can buffer temperatures as well as increase carbon storage (Mbow et
42 al. 2014a; Bioversity International 2016), and provide diversified and healthy diets in the face of
43 climate change.

44 CA adoption in Africa has been low despite more than three decades of implementation (Giller et al.
45 2009), although there is promising uptake recently in east and southern Africa. This calls for a better
46 understanding of the social and institutional aspects around CA adoption. Brown et al. (2017a) found

1 that institutional and community constraints hampered the use of financial, physical, human and
2 informational resources to implement CA programs.

3 Gender is another variable to consider since at intra-household level, decision-making and benefits
4 distribution. CA interventions have implications for labour requirements, labour allocation, and
5 investment decisions, all of which impacting the roles of men and women (Farnworth et al. 2016) (see
6 also Section 5.1.3). For example, in the global South, CA generally reduces labour and production
7 costs and generally leads to increased returns to family labour (Aryal et al. 2015) although a gender
8 shift of the labour burden to women have also been described (Giller et al. 2009).

9

10 **5.6.4.4 Sustainable intensification**

11 The need to produce about 50% more food by 2050, required to feed the increasing world population
12 (FAO 2018a) may come at the price of significant increases in GHG emissions and environmental
13 impacts, including loss of biodiversity. For instance, land conversion for agriculture is responsible for
14 an estimated 8-10% of all anthropogenic GHG emissions currently (see Section 5.4). Recent calls for
15 sustainable intensification (SI) are based on the premise that damage to the environment through
16 extensification outweighs benefits of extra food produced on new lands (Godfray 2015). However,
17 increasing net production area by restoring already degraded land may contribute to increased
18 production on the one hand and increase carbon sequestration on the other (Jat et al. 2016), thereby
19 contributing to both increased agricultural production and improved natural capital outcomes (Pretty
20 et al. 2018).

21 Sustainable intensification is a goal but does not specify *a priori* how it could be attained, e.g., which
22 agricultural techniques to deploy (Garnett et al. 2013). It can be combined with selected other
23 improved management practices, e.g., conservation agriculture (see above) or agroforestry, with
24 additional economic, ecosystem services, and carbon benefits. Sustainable intensification, by
25 improving nutrient-, water- and other input-use-efficiency, not only helps to close yield gaps and
26 contribute to food security (Garnett et al. 2013), but also reduces the loss of such production inputs
27 and associated emissions (Sapkota et al. 2017c; Wollenberg et al. 2016). Closing yield gaps is a way
28 to become more efficient in use of land per unit production. Currently, most regions in Africa and
29 South Asia have attained less than 40% of their potential crop production (Pradhan et al. 2015).
30 Integrated farming systems (e.g., mixed crop-livestock, crop-aquaculture) are strategies to produce
31 more products per unit land, which in regard to food security, becomes highly relevant.

32 Sustainable intensification acknowledges that enhanced productivity needs to be accompanied by
33 maintenance of other ecosystem services and enhanced resilience to shocks (Vanlauwe et al. 2014). SI
34 in intensively farmed areas may require a reduction in production in favour of increasing
35 sustainability in the broad sense (Buckwell et al. 2014) (see Cross-Chapter Box 6: Agricultural
36 Intensification). Hence, moving towards sustainability may imply lower yield growth rates than those
37 maximally attainable in such situations. For areas that contain valuable natural ecosystems, such as
38 the primary forest in the Congo basin, intensification of agriculture is one of the pillars of the strategy
39 to conserve forest (Vanlauwe et al. 2014). Intensification in agriculture is recognised as one of the
40 pathways to meet food security and climate change adaptation and mitigation goals (Sapkota et al.
41 2017c).

42 However, SI does not always confer co-benefits in terms of food security and climate change
43 adaption/mitigation. For example, in the case of Vietnam, intensified production of rice and pigs
44 reduced GHG emissions in the short term through land sparing, but after two decades, the emissions
45 associated with higher inputs were likely to outweigh the savings from land sparing (Thu Thuy et al.
46 2009). Intensification needs to be sustainable in all components of food system by curbing agricultural

1 sprawl, rebuilding soils, restoring degraded lands, reducing agricultural pollution, increasing water
2 use efficiency, and decreasing the use of external inputs (Cook et al, 2015).

3 A study conducted by Palm et al. (2010) in sub-Saharan Africa, reported that at low population
4 densities and high land availability, food security and climate mitigation goals can be met with
5 intensification scenarios, resulting in surplus crop area for reforestation. In contrast, for high
6 population density and small farm sizes, attaining food security and reducing GHG emissions require
7 use of more mineral fertilisers to make land available for reforestation. However, some forms of
8 intensification in drylands can increase rather than reduce vulnerability due to adverse effects such as
9 environmental degradation and increased social inequity (Robinson et al. 2015).

10 Sustainable intensification has been critiqued for considering food security only from the supply side,
11 whereas global food security requires attention to all aspects of food system, including access,
12 utilisation, and stability (Godfray 2015). Further, adoption of high-input forms of agriculture under
13 the guise of simultaneously improving yields and environmental performance will attract more
14 investment leading to higher rate of adoption but with the environmental component of SI quickly
15 abandoned (Godfray 2015). Where adopted, SI needs to engage with the sustainable development
16 agenda to (i) identify SI agricultural practices that strengthen rural communities, improve smallholder
17 livelihoods and employment, and avoid negative social and cultural impacts, including loss of land
18 tenure and forced migration; (ii) invest in the social, financial, natural, and physical capital needed to
19 facilitate SI implementation; and (iii) develop mechanisms to pay poor farmers for undertaking
20 sustainability measures (e.g., GHG emissions mitigation or biodiversity protection) that may carry
21 economic costs (Garnett et al. 2013).

22 In summary, integrated agricultural systems and practices can enhance food system resilience to
23 climate change and reduce GHG emissions, while helping to achieve sustainability (*high confidence*).

24

25 **Cross-Chapter Box 6: Agricultural intensification: land sparing, land** 26 **sharing and sustainability**

27 Eamon Haughey (Ireland), Tim Benton (United Kingdom), Annette Cowie (Australia), Lennart
28 Olsson (Sweden), Pete Smith (United Kingdom)

29 **Introduction**

30 The projected demand for more food, fuel and fibre for a growing human population necessitates
31 intensification of current land use to avoid conversion of additional land to agriculture and potentially
32 allow the sparing of land to provide other ecosystem services, including carbon sequestration,
33 production of biomass for energy, and the protection of biodiversity (Benton et al. 2018; Garnett et al.
34 2013). Land use intensity may be defined in terms of three components; (i) intensity of system inputs
35 (land/soil, capital, labour, knowledge, nutrients and other chemicals), (ii) intensity of system outputs
36 (yield per unit land area or per specific input) and (iii) the impacts of land use on ecosystem services
37 such as changes in soil carbon or biodiversity (Erb et al. 2013). Intensified land use can lead to
38 ecological damage as well as degradation of soil resulting in a loss of function which underpins many
39 ecosystem services (Wilhelm and Smith 2018); (Smith et al. 2016). Therefore, there is a risk that
40 increased agricultural intensification could deliver short-term production goals at the expense of
41 future productive potential, jeopardising long term food security (Tilman et al. 2011).

42 Agroecosystems which maintain or improve the natural and human capital and services they provide
43 may be defined as sustainable systems, while those which deplete these assets as unsustainable (Pretty
44 and Bharucha 2014). Producing more food, fuel and fibre without the conversion of additional non-
45 agricultural land while simultaneously reducing environmental impacts requires what has been termed

1 sustainable intensification (Godfray et al. 2010; FAO 2011e); see glossary and Cross-Chapter Box 6,
2 Figure 1). Sustainable intensification (SI) may be achieved through a wide variety of means; from
3 improved nutrient and water use efficiency via plant and animal breeding programs, to the
4 implementation of integrated soil fertility and pest management practices, as well as by smarter land
5 use allocation at a larger spatial scale: for example, matching land use to the context and specific
6 capabilities of the land (Benton et al. 2018). However, implementation of SI is broader than simply
7 increasing the technical efficiency of agriculture (“doing more with less”); it sometimes may require a
8 reduction of yields to raise sustainability, and successful implementation can be dependent on place
9 and scale. (Pretty et al. 2018), following (Hill 1985), highlights three elements to SI: (i) increasing
10 efficiency, (ii) substitution of less beneficial or efficient practices for better ones, and (iii) system
11 redesign to adopt new practices and farming systems (see Cross-Chapter Box 6, Table 1).

12 Under a land sparing strategy, intensification of land use in some areas, generating higher productivity
13 per unit area of land, can allow other land to provide other ecosystem services such as increased
14 carbon sequestration and the conservation of natural ecosystems and biodiversity (Balmford et al.
15 2018; Strassburg et al. 2014). Conversely under a land sharing strategy less, or no, land is set aside,
16 but lower levels of intensification are applied to agricultural land, providing a combination of
17 provisioning and other functions such as biodiversity conservation from the same land (Green et al.
18 2005). The two approaches are not mutually exclusive and the suitability of their application is
19 generally system-, scale- and/or location specific (Fischer et al. 2014). One crucial issue for the
20 success of a land sparing strategy is that spared land is protected from further conversion: as the
21 profits from the intensively managed land increase, there is an incentive for conversion of additional
22 land for production (Byerlee et al. 2014). Furthermore, it is implicit that there are limits to the SI of
23 land at a local and also planetary boundary level (Rockström et al. 2009). These may relate to the
24 “health” of soil, the presence of supporting services, such as pollination, local limits to water
25 availability, or limits on air quality. This implies that it may not be possible to meet demand
26 “sustainably” if demand exceeds local and global limits. There are no single global solutions to these
27 challenges and specific in situ responses for different farming systems and locations are required.
28 Bajželj et al. (2014) showed that implementation of SI, primarily through yield gap closure, had better
29 environmental outcomes compared with business as usual trajectories. However, SI alone will not be
30 able to deliver the necessary environmental outcomes from the food system – dietary change and
31 reduced food waste are also required (Springmann et al. 2018a; Bajželj et al. 2014).

32 **Cross-Chapter Box 6, Table 1 Approaches to sustainable intensification of agriculture (Pretty et al. 2018;**
33 **Hill 1985)**

| Approach | Sub-category | Examples/notes |
|-----------------------------|-------------------------|---|
| Improving efficiency | Precision agriculture | High and low-technology options to optimise resource use. |
| | Genetic improvements | Improved resource use efficiency through crop or livestock breeding. |
| | Irrigation technology | Increase production in areas currently limited by precipitation (sustainable water supply required). |
| | Organisational scale-up | Increasing farm organisational scale (e.g. co-operative schemes) can increase efficiency via facilitation of mechanisation and precision techniques. |
| Substitution | Green fertiliser | Replacing chemical fertiliser with green manures, compost (including vermicompost), biosolids and digestate (by product of anaerobic digestion) to maintain and improve soil fertility. |
| | Biological control | Pest control through encouraging natural predators. |
| | Alternative crops | Replacement of annual with perennial crops reducing the need for soil disturbance and reducing erosion. |
| | Premium products | Increase farm-level income for less output by producing a premium product. |
| System redesign | System diversification | Implementation of alternative farming systems: organic, agroforestry and intercropping (including the use of legumes). |
| | Pest management | Implementing integrated pest and weed management to reduce the quantities of inputs required. |
| | Nutrient management | Implementing integrated nutrient management by using crop and soil specific nutrient management – guided by soil testing. |
| | Knowledge transfer | Using knowledge sharing and technology platforms to accelerate the uptake of good agricultural practices. |

Improved efficiency – example of precision agriculture

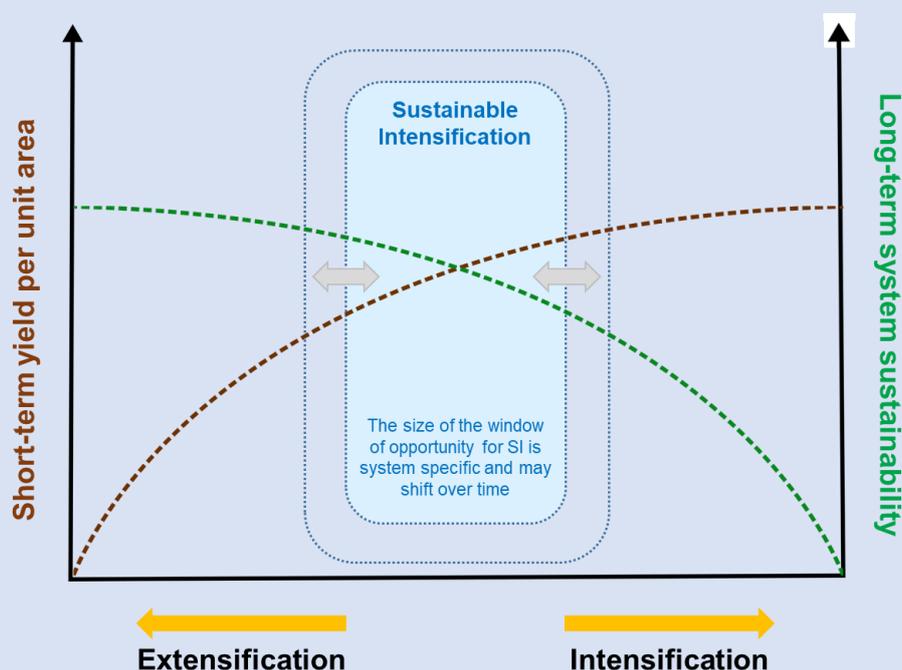
Precision farming usually refers to optimising production in fields through site-specific choices of crop varieties, agrochemical application, precise water management (e.g. in given areas or threshold moistures) and management of crops at a small scale (or livestock as individuals) (Hedley 2015). Precision agriculture has the potential to achieve higher yields in a more efficient and sustainable manner compared with traditional low-precision methods.

Precision agriculture

Precision agriculture is a technologically advanced approach that uses continual monitoring of crop and livestock performance to actively inform management practices. Precise monitoring of crop performance over the course of the growing season will enable farmers to economise on their inputs in terms of water, nutrients and pest management. Therefore, it can contribute to both the food security (by maintaining yields), sustainability (by reducing unnecessary inputs) and land sparing goals associated with SI. The site-specific management of weeds allows a more efficient application of herbicide to specific weed patches within crops (Jensen et al. 2012). Such precision weed control has resulted in herbicide savings of 19 – 22% for winter oilseed rape, 46 – 57% for sugar beet and 60 – 77% for winter wheat production (Gutjahr and Gerhards 2010). The use of on-farm sensors for real time management of crop and livestock performance can enhance farm efficiency (Aqeel-Ur-Rehman et al. 2014). Mapping soil nutrition status can allow for more targeted and therefore effective nutrient management practices (Hedley 2015). Using wireless sensors to monitor environmental conditions such as soil moisture, has the potential to allow more efficient crop irrigation (Srbnovska et al. 2015). Controlled traffic farming, where farm machinery is confined to permanent tracks, using automatic steering and satellite guidance, increases yields by minimising soil compaction. However, barriers to the uptake of many of these high-tech precision agriculture technologies remain. In what is described as the ‘implementation problem’, despite the potential to collect vast quantities of data on crop or livestock performance, applying these data to inform management decisions remains a challenge (Lindblom et al. 2017).

1 *Low-tech precision agriculture*

2 The principle of precision agriculture can be applied equally to low capital-input farming, in the form
 3 of low-tech precision agriculture (Conway 2013). The principle is the same but instead of adopting
 4 capital-heavy equipment (such as sensor technology connected to the ‘internet of things’, or large
 5 machinery and expensive inputs), farmers use knowledge and experience and innovative approaches
 6 often re-purposed, such as a bottle cap as a fertiliser measure for each plant, applied by hand (Mondal
 7 and Basu 2009). This type of precision agriculture is particularly relevant to small-scale farming in
 8 the global South, where capital investment is major limiting factor. For example, the application of a
 9 simple seed priming technique resulted in a 20 to 30% increase in yields of pearl millet and sorghum
 10 in semi-arid West Africa (Aune et al. 2017). Low-tech precision agriculture has the potential to
 11 increase the economic return per unit land area while also creating new employment opportunities.



14 **Cross-Chapter Box 6, Figure 1** There is a need to balance increasing demands for food, fuel and fibre
 15 with long-term sustainability of land use. Sustainable intensification can, in theory, offer a window of
 16 opportunity for the intensification of land use without causing degradation. This potentially allows the
 17 sparing of land to provide other ecosystem services, including carbon sequestration and the protection of
 18 biodiversity. However, the potential for SI is system specific and may change through time (indicated by
 19 grey arrows). Current practice may already be outside of this window and be unsustainable in terms of
 20 negative impacts on the long-term sustainability of the system

21 **Sustainable intensification through farming system redesign**

22 SI requires equal weight to be placed on the sustainability and intensification components (Benton
 23 2016; Garnett et al. 2013), Cross-Chapter Box 6, Figure 1 outlines the trade-offs which SI necessitates
 24 between the intensity of land use against long-term sustainability. One approach to this challenge is
 25 through farming system redesign including increased diversification.

26 *Diversification of intensively managed systems*

27 Incorporating higher levels of plant diversity in agroecosystems can improve the sustainability of
 28 farming systems (Isbell et al. 2017). Where intensive land use has led to land degradation more
 29 diverse land use systems such as intercropping can provide a more sustainable land use option with

1 co-benefits for food security, adaptation and mitigation objectives. For example, in temperate regions,
2 highly productive agricultural grasslands used to produce meat and dairy products are characterised
3 by monoculture pastures with high agrochemical inputs. Multi-species grasslands may provide a route
4 to SI, as even a modest increase in species richness in intensively managed grasslands can result in
5 higher forage yields without increased inputs, such as chemical fertiliser (Finn et al. 2013; Sanderson
6 et al. 2013; Tilman et al. 2011). Recent evidence also indicates multispecies grasslands have greater
7 resilience to drought, indicating co-benefits for adaptation (Hofer et al. 2016; Haughey et al. 2018).

8 *Diversification of production systems*

9 Agroforestry systems (see glossary) can promote regional food security and provide many additional
10 ecosystem services when compared with monoculture crop systems. Co-benefits for mitigation and
11 adaptation include increased carbon sequestration in soils and biomass, improved water and nutrient
12 use efficiency and the creation of favourable micro-climates (Waldron et al. 2017). Silvopasture
13 systems, which combine grazing of livestock and forestry, are particularly useful in reducing land
14 degradation where the risk of soil erosion is high (Murgueitio et al. 2011). Crop and livestock systems
15 can also be combined to provide multiple services. Perennial wheat derivatives produced both high
16 quality forage and substantial volumes of cereal grains (Newell and Hayes 2017), and show promise
17 for integrating cereal and livestock production while sequestering soil carbon (Ryan et al. 2018). A
18 key feature of diverse production systems is the provision of multiple income streams for farming
19 households, providing much needed economic resilience in the face of fluctuation of crop yields and
20 prices.

21 **Landscape Approaches**

22 The land sparing and land sharing approaches which may be used to implement SI are inherently
23 “landscape approaches” (e.g., (Hodgson et al. 2010)). While the term landscape is by no means
24 precise (Englund et al. 2017), landscape approaches, focused for example at catchment scale, are
25 generally agreed to be the best way to tackle competing demands for land (e.g. Sayer et al. 2013), and
26 are the appropriate scale at which to focus the implementation of sustainable intensification. The
27 landscape approach allots land to various uses – cropping, intensive and extensive grazing, forestry,
28 mining, conservation, recreation, urban, industry, infrastructure – through a planning process that
29 seeks to balance conservation and production objectives. With respect to SI, a landscape approach is
30 pertinent to achieving potential benefits for biodiversity conservation, ensuring that land “spared”
31 through SI remains protected, and that adverse impacts of agriculture on conservation land are
32 minimised. Depending on the land governance mechanisms applied in the jurisdiction, different
33 approaches will be appropriate/required. However, benefits are only assured if land use restrictions
34 are devised and enforced.

35 **Summary**

36 Intensification needs to be achieved sustainably, necessitating a balance between productivity today
37 and future potential (*high agreement, medium evidence*). Improving efficiency of agriculture systems
38 can increase production per unit of land through greater resource use efficiency. To achieve SI some
39 intensively managed agricultural systems may have to be diversified as they cannot be further
40 intensified without land degradation. A combination of land sparing and sharing options can be
41 utilised to achieve SI – their application is most likely to succeed if applied using a landscape
42 approach.

44 **5.6.5 Role of urban agriculture**

45 Cities are an important actor in the food system in regard to both demand for food for urban dwellers
46 and production of food in urban and peri-urban areas (see also cross-chapter box 4: Climate Change

1 and Urbanisation in Chapter 2). Both the demand side and supply side roles are important relative to
2 climate change mitigation and adaptation strategies. Urban areas are home to more than half of the
3 world's population, and a minimal proportion of the production; thus, they are important drivers for
4 the development of the complex food systems in place today, in regard to supply chains and dietary
5 preferences.

6 The increasing separation of urban and rural populations with regard to territory and culture is one of
7 the factors favouring the nutrition transition towards urban diets (Weber and Matthews 2008; Neira et
8 al. 2016). These are primarily based on a high diversity of food products, independent of season and
9 local production, and on the extension of the distances that food travels between production and
10 consumption. The transition of traditional diets to more homogeneous diets has also become tied to
11 consumption of animal protein, which has increased GHG emissions globally (see also Section 5.4.6).

12 Cities are becoming key actors in developing strategies of mitigation to climate change, in their food
13 procurement and in sustainable urban food policies alike (McPhearson et al. 2018). These are being
14 developed by big and medium-sized cities in the world, often integrated within climate change
15 policies (Moragues et al. 2013; Calori and Magarini 2015). A review conducted of 100 cities across
16 the world shows that urban food consumption is one of the largest sources of urban material flows,
17 urban carbon footprint, and land footprint (Goldstein et al. 2017). Additionally, the urban poor have
18 limited capacity to adapt to climate-related impacts, which place their food security at risk under
19 climate change (Dubbeling and de Zeeuw 2011).

20 *Urban and peri-urban areas.* In 2010, around 14% of the global population was nourished by food
21 grown in urban and peri-urban areas (Kriewald et al.). A review study on sub-Saharan Africa shows
22 that urban and peri-urban agriculture contributes to climate change adaptation and mitigation (Lwasa
23 et al. 2014, 2015). Urban and peri-urban agriculture reduces food carbon footprint by avoiding long
24 distance food transport and limits GHG emissions by recycling organic waste and wastewater that
25 would otherwise releases methane from landfill and dumping sites (Lwasa et al. 2014). Urban and
26 peri-urban agriculture also contributes in adapting to climate change including extreme events, by
27 reducing urban heat island effect, increasing water infiltration and slowing down run-offs to prevent
28 flooding, etc. (Lwasa et al. 2014, 2015; Kumar et al. 2017a). For example, a scenario analysis shows
29 that urban gardens reduce the surface temperature up to 10°C in comparison to the temperature
30 without vegetation (Tsilini et al. 2015). Urban agriculture can also improve biodiversity and
31 strengthen associated ecosystem services (Lin et al. 2015).

32 Urban and peri-urban agriculture is exposed to climate risks and urban growth that may undermine its
33 long-term potential to address urban food security (Padgham et al. 2015). Therefore, there is a need to
34 better understand the impact of urban sprawl on peri-urban agriculture; the contribution of urban and
35 peri-urban agriculture to food self-sufficiency of cities; the risks posed by pollutants from urban areas
36 to agriculture and vice-versa; the global and regional extent of urban agriculture; and the role that
37 urban agriculture could play in climate resilience and abating malnutrition (Mok et al. 2014; Hamilton
38 et al. 2014). Globally, urban sprawl is projected to consume 1.8–2.4% and 5% of the current
39 cultivated land by 2030 and 2050 respectively, leading to crop calorie loss of 3–4% and 6–7%,
40 respectively (Pradhan et al. 2014; Bren d'Amour et al. 2017). Kriewald et al. shows that the urban
41 growth has the largest impacts in most of the sub-continent (e.g., Western, Middle, and Eastern
42 Africa) while climate change will mostly reduce potential of urban and peri-urban agriculture in
43 Southern Europe and Northern Africa.

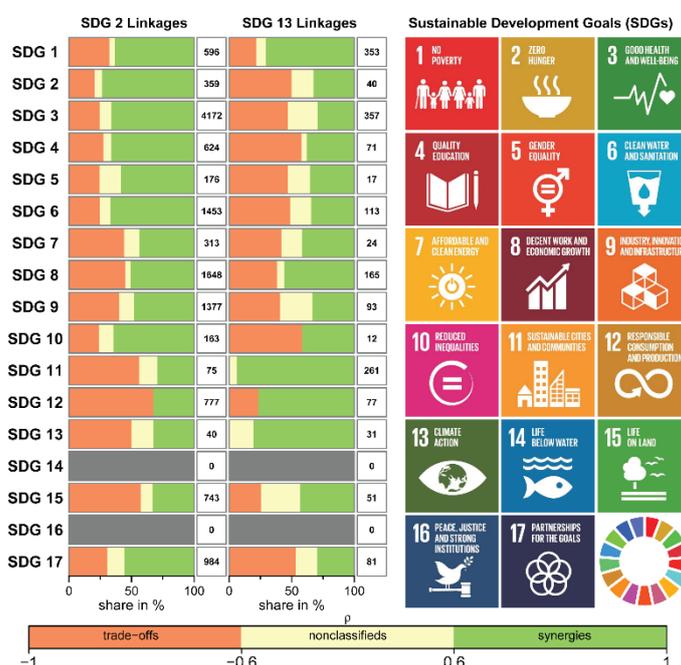
44 In summary, urban and peri-urban agriculture can contribute to improving urban food security,
45 reducing greenhouse gas emissions, and adapting to climate change impacts (*robust evidence, medium
46 agreement*).

47

1 5.6.6 Links to the Sustainable Development Goals

2 In 2015, the Sustainable Development Goals (SDGs) and the Paris Agreement were two global major
3 international policies adopted by all countries to guide the world to overall sustainability, within the
4 2030 Sustainable Development Agenda and UNFCCC processes respectively. The 2030 Sustainable
5 Development agenda includes 17 goals and 169 targets, including zero hunger, sustainable agriculture
6 and climate action (United Nations 2015).

7 This section focuses on intra- and inter-linkages of SDG 2 and SDG 13 based on the official SDG
8 indicators (Figure 5.16), showing the current conditions (see (Roy et al. 2018) and Chapter 7 for
9 further discussion). The second goal (Zero Hunger – SDG 2) aims to end hunger and all forms of
10 malnutrition by 2030 and commits to universal access to safe, nutritious and sufficient food at all
11 times of the year. SDG 13 (Climate Action) calls for urgent action to combat climate change and its
12 impacts. Integrating the SDGs into the global food system can provide opportunities for mitigation
13 and adaptation and enhancement of food security.



14
15 **Figure 5.16. Intra and inter-linkages for SDG 2 (*Zero hunger*) and SDG 13 (*Climate action*) at the global**
16 **level using the official indicators of Sustainable Development Goals that consists data for 122 indicators**
17 **for a total of 227 countries between the years 1983 and 2016 (United Nations Statistics Division 2016) and**
18 **applying a statistical approach (Pradhan et al. 2017). Pradhan et al. (2017) defined synergy and trade-offs**
19 **as significant positive ($\rho > 0.6$, red bar) and negative ($\rho < -0.6$, green bar) spearman correlation between**
20 **SDG indicators, respectively. The ρ between 0.6 and -0.6 is considered as nonclassifieds (yellow bar). The**
21 **correlation between unique pairs of indicator time-series is carried based on country data, e.g., between**
22 **“prevalence of undernourishment” (an indicator for SDG 2.1) and “maternalmortality ratio” (an**
23 **indicator for SDG 3.1). The data pairs can belong to the same goal or to two distinct goals. At the global**
24 **level, intra-linkages of SDGs are quantified by the percentage of synergies, trade-offs, and nonclassifieds**
25 **of indicator pairs belonging to the same SDG (here, SDG 2 and SDG 13) for all the countries. Similarly,**
26 **SDG interlinkages are estimated by the percentage of synergies, trade-offs, and nonclassifieds between**
27 **indicator pairs that fall into two distinct goals for all the countries. The grey bar shows insufficient data**
28 **for analysis. The number of data pair used for the analysis is presented in the grey box.**

29 Ensuring food security (SDG 2) shows positive relations (synergies) with most goals (Pradhan et al.
30 2017; International Council for Science (ICSU) 2017), but has trade-offs with SDG 12 (*Responsible*
31 *Consumption and Production*) and SDG 15 (*Life on Land*) under current development paradigms

1 (Pradhan et al. 2017). Sustainable transformation of traditional consumption and production
2 approaches can overcome these trade-offs based on several innovative methods (Shove et al. 2012).
3 For example, sustainable intensification and reduction of food waste can minimise the observed
4 negative relations between SDG 2 and other goals (Obersteiner et al. 2016) (see also Cross-Chapter
5 Box 6 and Section 5.5.2). Achieving the target 12.3 of SDG 12 (Responsible Consumption and
6 Production) “by 2030, to halve per capita global food waste at the retail and consumer levels and
7 reduce food losses along production and supply chains, including post-harvest losses” will contribute
8 to climate change mitigation.

9 Doubling productivity of smallholder farmers and halving food loss and waste by 2030 are targets of
10 SDG 2 and SDG 12, respectively (United Nations Statistics Division 2016). Agroforestry that
11 promotes biodiversity and sustainable land management also contributes to food security (Montagnini
12 and Metzler 2017). Land restoration and protection (SDG 15) can increase crop productivity (SDG 2)
13 (Wolff et al. 2018). Similarly, efficient irrigation practices can reduce water demand for agriculture
14 that could improve health of the freshwater ecosystem (SDG 6 and SDG 15) without reducing food
15 production (Jägermeyr et al. 2017).

16 Climate action (SDG 13) shows negative relations (trade-offs) with most goals and antagonistic to the
17 2030 development agenda under the current development paradigm (Figure 5.16) (Lusseau and
18 Mancini 2019; Pradhan 2019). The targets for SDG 13 have a high focus on climate change
19 adaptation and the data for the SDG 13 indicators are limitedly available. SDG 13 shares two
20 indicators with SDG 1 and SDG 11 (United Nations 2017) and therefore, has mainly positive linkages
21 with these two goals. Trade-offs was observed between SDG 2 and SDG 13 for around 50% of the
22 case (Pradhan et al. 2017).

23 Transformation from current development paradigms and breaking of these lock-in effects can protect
24 climate and achieve food security in future. Sustainable agriculture practices can provide climate
25 change adaptation and mitigation synergies, linking SDG 2 and SDG 13 more positively
26 (International Council for Science (ICSU) 2017). IPCC highlights that most of the current observed
27 trade-offs between SDG 13 and other SDGs can be converted into synergies based on various
28 mitigation options that can be deployed to limit the global warming well below 1.5°C (IPCC 2018b).

29 In summary, there are fundamental synergies that can facilitate the joint implementation of strategies
30 to achieve SDGs and climate action, with particular reference to those climate response strategies
31 related to both supply side (production and supply chains) and demand side (consumption and dietary
32 choices) described in this chapter (*high agreement and medium evidence*).

34 **5.7 Enabling conditions and knowledge gaps**

35 To achieve mitigation and adaptation to climate change in food systems, enabling conditions are
36 needed to scale up the adoption of effective strategies (such as those described in Sections 5.3 to 5.6
37 and Chapter 6). These enabling conditions include multi-level governance and multi-sector
38 institutions (Supplementary Material Section SM5.7) and multiple policy pathways (Section 5.7.1,
39 5.7.2). In this regard, the subnational level is gaining relevance both in food systems and climate
40 change. Just Transitions are needed to address both climate change and food security (Section 5.7.3).
41 Mobilisation of knowledge, education, and capacity will be required (Section 5.7.4) to fill knowledge
42 gaps (Section 5.7.5).

43 Effective governance of food systems and climate change requires the establishment of institutions
44 responsible for coordinating among multiple sectors (education, agriculture, environment, welfare,
45 consumption, economic, health), levels (local, regional, national, global) and actors (governments,
46 CSO, public sector, private sector, international bodies). Positive outcomes will be enegendered by

1 participation, learning, flexibility, and cooperation. See Supplementary Material Section SM5.7 for
2 further discussion.

3

4 **5.7.1 Enabling policy environments**

5 The scope for responses to make sustainable land use inclusive of climate change mitigation and
6 adaptation, and the policies to implement them, are covered in detail in Chapters 6 and 7. Here we
7 highlight some of the major policy areas that have shaped the food system, and might be able to shape
8 responses in future. Although two families of policy – agriculture and trade – have been instrumental
9 in shaping the food system in the past (and potentially have led to conditions that increase climate
10 vulnerability) (Benton and Bailey 2019), a much wider family of policy instruments can be deployed
11 to reconfigure the food system to deliver healthy diets in a sustainable way.

12

13 **5.7.1.1 Agriculture and trade policy**

14 *Agriculture.* The thrust of agricultural policies over the last 50 years has been to increase productivity,
15 even if at the expense of environmental sustainability (Benton and Bailey 2019). For example, in
16 2007-9 46% of OECD support for agriculture was based on measures of output (price support or
17 payments based on yields), 37% of support was based on the current or historical area planted, herd
18 size (or correlated measures of the notional costs of farming), and 13% was payments linked to input
19 prices. In a similar vein, non-OECD countries have promoted productivity growth for their
20 agricultural sectors.

21 *Trade.* Along with agricultural policy to grow productivity, the development of frameworks to
22 liberalise trade (such as the General Agreement on Tariffs and Trade (GATT) Uruguay Round, now
23 incorporated into the World Trade Organisation) have been essential in stimulating the growth of a
24 globalised food system. Almost every country has a reliance on trade to fulfil some or all of its local
25 food needs, and trade networks have grown to be highly complex (Puma et al. 2015; MacDonald et al.
26 2015; Fader et al. 2013; Ercsey-Ravasz et al. 2012). This is because many countries lack the capacity
27 to produce sufficient food due to climatic conditions, soil quality, water constraints, and availability of
28 farmland (FAO 2015b). In a world of liberalised trade, using comparative advantage to maximise
29 production in high-yielding commodities, exporting excess production, and importing supplies of
30 other goods supports economic growth.

31 City states as well as many small island states, do not have adequate farmland to feed their
32 populations, while sub-Saharan African countries are projected experience high population growth as
33 well as to be negatively impacted by climate change, and thus will likely find it difficult to produce all
34 of their own food supplies (Agarwal et al. 2002). One study estimates that some 66 countries are
35 currently incapable of being self-sufficient in food (Pradhan et al. 2014). Estimates of the proportion
36 of people relying on trade for basic food security vary from ~16% to ~22% (Fader et al. 2013;
37 Pradhan et al. 2014), with this figure rising to between 1.5 and 6 billion people by 2050, depending on
38 dietary shifts, agricultural gains, and climate impacts (Pradhan et al. 2014).

39 Global trade is therefore essential for achieving food and nutrition security under climate change
40 because it provides a mechanism for enhancing the efficiency of supply chains, reducing the
41 vulnerability of food availability to changes in local weather, and moving production from areas of
42 surplus to areas of deficit (FAO 2018d). However, the benefits of trade will only be realised if trade is
43 managed in ways that maximise broadened access to new markets while minimising the risks of
44 increased exposure to international competition and market volatility (Challinor et al. 2018; Brown et
45 al. 2017b).

1 As described in Section 5.8.1, trade acts to buffer exposure to climate risks when the market works
 2 well. Under certain conditions – such as shocks, or the perception of a shock, coupled with a lack of
 3 food stocks or lack of transparency about stocks (Challinor et al. 2018; Marchand et al. 2016) – the
 4 market can fail and trade can expose countries to food price shocks.

5 Furthermore, Clapp (2016) showed that trade, often supported by high levels of subsidy support to
 6 agriculture in some countries, can depress world prices and reduce incomes for other agricultural
 7 exporters. Lower food prices that result from subsidy support may benefit urban consumers in
 8 importing countries, but at the same time they may hurt farmers' incomes in those same countries. The
 9 outmigration of smallholder farmers from the agriculture sector across the Global South is
 10 significantly attributed to these trade patterns of cheap food imports (Wittman 2011; McMichael
 11 2014; Akram-Lodhi and others 2013). Food production and trade cartels, as well as financial
 12 speculation on food futures markets, affect low-income market-dependent populations.

13 Food sovereignty is a framing developed to conceptualise these issues (Reuter 2015). They directly
 14 relate to the ability of local communities and nations to build their food systems, based among other
 15 aspects, on diversified crops and indigenous and local knowledge. If a country enters international
 16 markets by growing more commodity crops and reducing local crop varieties, it may get economic
 17 benefits, but may also expose itself to climate risks and food insecurity by increasing reliance on
 18 trade, which may be increasingly disrupted by climate risks. These include a local lack of resilience
 19 from reduced diversity of products, but also exposure to food price spikes, which can become
 20 amplified by market mechanisms such as speculation.

21 In summary, countries must determine the balance between locally produced vs imported food (and
 22 feed) such that it both minimises climate risks and ensures sustainable food security. There is *medium*
 23 *evidence* that trade has positive benefits but also creates exposure to risks (Section 5.3).

24

25 **5.7.1.2 Scope for expanded policies**

26 There are a range of ways that policy can intervene to stimulate change in the food system – through
 27 agriculture, research and development, food standards, manufacture and storage, changing the food
 28 environment and access to food, changing practices to encourage or discourage trade (Table 5.6).
 29 Novel incentives can stimulate the market, for example, through reduction in waste or changes in diets
 30 to gain benefits from a health or sustainability direction. Different contexts with different needs will
 31 require different set of policies at local, regional and national levels. See Supplementary Material
 32 Section SM5.7 for further discussion on expanded policies.

33

34 **Table 5.6 Potential policy “families” for food-related adaptation and mitigation of climate change. The**
 35 **column “scale” refers to scale of implementation: International (I), national (N), sub-national-regional**
 36 **(R), and local (L).**

| Family | Sub-family | Scale | Interventions | Examples |
|-------------------------------|---|-------|----------------------------------|---|
| Supply-side efficiency | Increasing agricultural efficiency and yields | I, N | Agricultural R&D | Investment in research, innovation, knowledge exchange, e.g., on genetics, yield gaps, resilience |
| | | I, N | Supporting precision agriculture | Agricultural engineering, robotics, big data, remote sensing, inputs |
| | | I, N | Sustainable intensification | Soils, nutrients, capital, labour (see |

| | | | | |
|--|---------------------------------------|---------|---|--|
| | | | projects | Cross-Chapter Box 6) |
| | | N, R | Improving farmer training and knowledge sharing | Extension services, online access, field schools, farmer-to-farmer networks (CABI 2019) |
| | Land use planning | N, R, L | Land use planning for ecosystem services (remote sensing, indigenous and local knowledge) | Zoning, protected area networks, multifunctional landscapes, “land sparing” (see Cross-Chapter Box 6; Benton et al. 2018; Jones et al. 2013) |
| | | N, R, L | Conservation agriculture programs | Soil and water erosion control, soil quality improvement (Conservation Evidence 2019) |
| | | N | Payment for ecosystem services | Incentives for farmers/landowners to choose lower-profit but environmentally benign resource use, e.g., Los Negros Valley in Bolivia (Ezzine-de-Blas et al. 2016) |
| | Market approaches | I, N | Mandated carbon cost reporting in supply chains; public/private incentivised insurance products | Carbon and natural capital accounts (CDP 2019), crop insurance (Müller et al. 2017a) |
| | Trade | I | Liberalising trade flows; green trade | Reduction in GHG emissions from supply chains (Neumayer 2001) |
| Raising profitability and quality | Stimulating markets for premium goods | N, R | Sustainable farming standards, agroecology projects, local food movements | Regional policy development, public procurement of sustainable food (Mairie de Paris 2015) |
| Modifying demand | Reducing food waste | I, N, L | Regulations, taxes | ‘Pay-As-You-Throw (PAYT)’ schemes; EU Landfill Directives; Japan Food Waste Recycling Law 2008; South Africa Draft Waste Classification and Management Regulations 2010 (Chalak et al. 2016) |
| | | I, N, L | Awareness campaigns, education | FAO Global Initiative on Food Loss and Waste Reduction (FAO 2019b) |
| | | I, N | Funding for reducing food waste | Research and investment for shelf life, processing, packaging, cold storage (MOFPI 2019) |
| | | I, N, L | Circular economy using waste as inputs | Biofuels, distribution of excess food to charities (Baglioni et al. 2017) |
| | Reducing consumption | I, N, L | Carbon pricing for selected food commodities | Food prices reflective of GHG gas emissions throughout production and |

| | | | | |
|--|--|---------|---|--|
| | n of carbon-intensive food | | | supply chain (Springmann et al. 2017; Hasegawa et al. 2018) |
| | | I, N, L | Changing food choice through education | Nutritional and portion-size labelling, ‘nudge’ strategies (positive reinforcement, indirect suggestion) (Arno and Thomas 2016) |
| | | I, N, L | Changing food choices through money transfers | Unconditional cash transfers; e-vouchers exchanged for set quantity or value of specific, pre-selected goods (Fenn 2018) |
| | | N, L | Changing food environments through planning | Farmers markets, community food production, addressing ‘food deserts’ (Ross et al. 2014) |
| | Combining carbon and health objectives | I, N, L | Changing subsidies, standards, regulations to healthier and more sustainably produced foods | USDA’s “Smart Snacks for School” regulation mandating nutritional guidelines (USDA 2016) Incentivising production via subsidies (direct to producer based on output or indirect via subsidising inputs) |
| | | N | Preventative vs curative public health care incentives | Health insurance cost reductions for healthy and sustainable diets |
| | | I, N, L | Food system labelling | Organic certification, nutrition labels, blockchain ledgers (Chadwick 2017) |
| | | N, L | Education and awareness campaigns | School curricula; public awareness campaigns |
| | | N, L | Investment in disruptive technologies (e.g., cultured meat) | Tax breaks for R&D, industrial strategies (European Union 2018) |
| | | N, L | Public procurement | For health: Public Procurement of Food for Health (Caldeira et al. 2017) For environment: Paris Sustainable Food Plan 2015-2020 Public Procurement Code (Mairie de Paris 2015) |

1

2 In summary, although agriculture is often thought to be shaped predominantly by agriculture and trade
3 policies, there are over twenty families of policy areas that can shape agricultural production directly
4 or indirectly (through environmental regulations or through markets, including by shaping consumer

1 behaviour). Thus, delivering outcomes promoting climate change adaptation and mitigation can arise
2 from policies across many departments, if suitably designed and aligned.

4 **5.7.1.3 Health-related policies and cost savings**

5 The co-benefits arising from mitigating climate change through changing dietary patterns, and thus
6 demand, have potentially important economic impacts (*high confidence*). The gross value added from
7 agriculture to the global economy (GVA) was USD 1.9tn (in 2013 (FAO 2015c)), from a global
8 agriculture economy (GDP) of USD 2.7tn (in 2016). In 2013, the FAO estimated an annual cost of
9 USD 3.5tn for malnutrition (FAO 2013a).

10 However, this is likely to be an underestimate of the economic health costs of current food systems
11 for several reasons: (1) Lack of data – for example there is little robust data in the UK on the
12 prevalence of malnutrition in the general population (beyond estimates of obesity and surveys of
13 malnourishment of patients in hospital and care homes, from which estimates over 3 million people in
14 the UK are undernourished (BAPEN 2012)); (2) Lack of robust methodology to determine, for
15 example, the exact relationship between overconsumption of poor diets, obesity and non-
16 communicable diseases like diabetes, cardio-vascular disease, a range of cancers or Alzheimer's
17 disease (Peditizi et al. 2016), (3) Unequal healthcare spending around the world.

18 In the US, the economic cost of diabetes, a disease strongly associated with obesity and affecting
19 about 23 million Americans, is estimated at USD 327bn in 2017 (American Diabetes Association
20 2018), with direct healthcare costs of USD 9,600 per person. By 2025, it is estimated that globally
21 there will be over 700 million people with diabetes (NCD-RisC 2016b), over 30 times the number in
22 the US. Even if a global average cost of diabetes per capita were a quarter of that in the US, the total
23 economic cost of diabetes would be approximately the same as global agricultural GDP. Finally, (4)
24 the role of agriculture in causing ill-health beyond dietary health, such as through degrading air
25 quality (e.g., (Paulot and Jacob 2014)).

26 Whilst data of the healthcare costs associated with the food system and diets are scattered and the
27 proportion of costs directly attributable to diets and food consumption is uncertain, there is potential
28 for more preventative healthcare systems to save significant costs that could incentivise agricultural
29 business models to change what is grown, and how. The potential of moving towards more
30 preventative healthcare is widely discussed in the health economics literature, particularly in order to
31 reduce the life-style-related (including dietary-related) disease component in aging populations (e.g.,
32 (Bloom et al. 2015)).

34 **5.7.1.4 Multiple policy pathways**

35 As discussed in more detail in Chapters 6 and 7, there is a wide potential suite of interventions and
36 policies that can potentially enhance the adaptation of food systems to climate change, as well as
37 enhance the mitigation potential of food systems on climate change. There is an increasing number of
38 studies that argue that the key to sustainable land management is not in land management practices
39 but in the factors that determine the demand for products from land (such as food). Public health
40 policy therefore has the potential to affect dietary choice and thus the demand for different amounts
41 of, and types of, food.

42 Obersteiner et al. (2016) show that increasing the average price of food is an important policy lever
43 that, by reducing demand, reduces food waste, pressure on land and water, impacts on biodiversity
44 and through reducing emissions, mitigates climate change and potentially helps to achieve multiple
45 SDGs. Whilst such policy responses – such as a carbon tax applied to goods including food – has the
46 potential to be regressive, affecting the poor differentially (Frank et al. 2017; Hasegawa et al. 2018;

1 Kehlbacher et al. 2016), and increasing food insecurity – further development of social safety nets can
2 help to avoid the regressive nature (Hasegawa et al. 2018). Hasegawa et al. (2018) point out that such
3 safety nets for vulnerable populations could be funded from the revenues arising from a carbon tax.

4 The evidence suggests, as with SR15 (IPCC 2018a) and its multiple pathways to climate change
5 solutions, that there is no single solution that will address the problems of food and climate change,
6 but instead there is a need to deploy many solutions simultaneously adapted to the needs and options
7 available in a given context. For example, Springmann et al. (2018a) indicate that maintaining the
8 food system within planetary boundaries at mid-century, including equitable climate, requires
9 increasing the production (and resilience) of agricultural outputs (i.e., closing yield gaps), reducing
10 waste, and changes in diets towards ones often described as flexitarian (low-meat dietary patterns that
11 are in line with available evidence on healthy eating). Such changes can have significant co-benefits
12 for public health, as well as facing significant challenges to ensure equity (in terms of affordability for
13 those in poverty).

14 Significant changes in the food system require them to be acceptable to the public (“public license”),
15 or they will be rejected. Focus groups with members of the public around the world, on the issue of
16 changing diets, have shown that there is a general belief that the government plays a key role in
17 leading efforts for change in consumption patterns (Wellesley et al. 2015). If governments are not
18 leading on an issue, or indicating the need for it through leading public dialogue, it signals to their
19 citizens that the issue is unimportant or undeserving of concern

20 In summary, there is significant potential (*high confidence*) that, through aligning multiple policy
21 goals, multiple benefits can be realised that positively impact public health, mitigation and adaptation
22 (e.g. adoption of healthier diets, reduction in waste, reduction in environmental impact). These
23 benefits may not occur without the alignment across multiple policy areas (*high confidence*).

25 **5.7.2 Enablers for changing markets and trade**

26 “Demand” for food is not an exogenous variable to the food system but is shaped crucially by its
27 ability to produce, market, and supply food of different types and prices. These market dynamics can
28 be influenced by a variety of factors beyond consumer preferences (e.g., corporate power and
29 marketing, transparency, the food environment more generally), and the ability to reshape the market
30 can also depend on its internal resilience and/or external shocks (Challinor et al. 2018; Oliver et al.
31 2018)).

33 **5.7.2.1 Capital markets**

34 Two areas are often discussed in regard to role of capital markets in shaping the food system. First,
35 investment in disruptive technologies might stimulate climate-smart food systems (WEF/McKinsey &
36 Company 2018; Bailey and Wellesley 2017), including alternative proteins, such as laboratory or
37 “clean meat” (which has significant ability to impact on land use requirements) (Alexander et al.
38 2017) (See Section 5.5.1.6). An innovation environment through which disruptive technology can
39 emerge typically requires the support of public policy, whether in directly financing small and
40 emerging enterprises, or funding research and development via reducing tax burdens.

41 Second, widespread adoption of (and perhaps underpinned by regulation for) natural capital
42 accounting as well as financial accounting are needed. Investors can then be aware of the risk
43 exposure of institutions, which can undermine sustainability through externalising costs onto the
44 environment. The prime example of this in the realm of climate change is the Carbon Disclosure
45 Project, with around 2500 companies voluntarily disclosing their carbon footprint, representing nearly
46 60% of the world’s market capital (CDP 2018).

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46

5.7.2.2 Insurance and re-insurance

The insurance industry can incentivise actors' behaviour towards greater climate mitigation or adaptation, including building resilience. For example, Lloyd's of London analysed the implications of extreme weather for the insurance market, and conclude that the insurance industry needs to examine their exposure to risks through the food supply chain and develop innovative risk-sharing products can make an important contribution to resilience of the global food system (Lloyd 2015).

Many of these potential areas for enabling healthy and sustainable food systems are also knowledge gaps, in that whilst the levers are widely known, their efficacy and the ability to scale-up, in any given context, are poorly understood.

5.7.3 Just transitions to sustainability

Research is limited on how land use transitions would proceed from ruminant production to other socio-ecological farming systems. Ruminants have been associated with humans since the early development of agriculture, and the role of ruminants in many agricultural systems and smallholder communities is substantial. Ruminant production systems have been adapted to a wide range of socioeconomic and environmental conditions in crop, forestry, and food processing settings (Čolović et al. 2019), bioenergy production (de Souza et al. 2019), and food waste recycling (Westendorf 2000). Pasture cultivation in succession to crops is recognised as important to management of pest and diseases cycles and to improve soil carbon stocks and soil quality (Carvalho and Dedieu 2014). Grazing livestock is important as a reserve of food and economic stocks for some smallholders (Ouma et al. 2003).

Possible land-use options for transitions away from livestock production in a range of systems include (a) retain land but reduce investments to run a more extensive production system; (b) change land use by adopting a different production activity; (c) abandon land (or part of the farm) to allow secondary vegetation regrowth (Carvalho et al. 2019; Laue and Arima 2016); and (d) invest in afforestation or reforestation (Baynes et al. 2017). The extensification option could lead to increases rather than decreases in GHG emissions related to reduction in beef consumption. Large-scale abandonment, afforestation, or reforestation would probably have more positive environmental outcomes, but could result in economic and social issues that would require governmental subsidies to avoid decline and migration in some regions (Henderson et al. 2018). Alternative economic use of land, such as bioenergy production, could balance the negative socioeconomic impact of reducing beef output, reduce the tax values needed to reduce consumption, and avoid extensification of ruminant production systems (Wirsenius et al. 2011). However, the analysis of the transition of land use for ruminants to other agricultural production systems is still a literature gap (Cross-Chapter Box 7: Bioenergy and BECCS in mitigation scenarios, in Chapter 6).

Finally, it is important to recognise that, while energy alternatives produce the same function for the consumer, it is questionable that providing the same nutritional value through an optimised mix of dietary ingredients provides the same utility for humans. Food has a central role in human pleasure, socialisation, cultural identity, and health (Röös et al. 2017), including some of the most vulnerable groups, so just transitions and their costs need to be taken into account. Pilot projects are important to provide greater insights for large-scale policy design, implementation, and enforcement.

In summary, more research is needed on how land use transitions would proceed from ruminant production to other farming systems and affect the farmers and other food system actors involved. There is *limited evidence* on what the decisions of farmers under lower beef demand would be.

1 **5.7.4 Mobilising knowledge**

2 Addressing climate change-related challenges and ensuring food security requires all types of
3 knowledge (formal/non-formal, scientific/indigenous, women, youth, technological). Miles et al.
4 (2017) stated that a research and policy feedback that allows transitions to sustainable food systems
5 must have at first a whole system approach. Currently, in transmitting knowledge for food security
6 and land sustainability under climate change there are three major approaches: (1) public technology
7 transfer with demonstration (extension agents); (2) public and private advisory services (for
8 intensification techniques) and; (3) non-formal education with many different variants such as farmers
9 field schools, rural resource centers; facilitation extension where front-line agents primarily work as
10 “knowledge brokers” in facilitating the teaching–learning process among all types of farmers
11 (including women and rural young people), or farmer-to-farmer, where farmers act themselves as
12 knowledge transfer and sharing actors through peer processes.

13

14 **5.7.4.1 Indigenous and local knowledge**

15 Recent discourse has a strong orientation towards scaling-up innovation and adoption by local
16 farmers. However, autonomous adaptation, indigenous knowledge and local knowledge are both
17 important for agricultural adaptation (Biggs et al. 2013) (See Section 5.3). These involve the
18 promotion of farmer participation in governance structures, research, and the design of systems for the
19 generation and dissemination of knowledge and technology, so that farmers needs and knowledge can
20 be taken into consideration. Klenk et al. (2017) found that mobilisation of local knowledge can inform
21 adaptation decision-making and may facilitate greater flexibility in government-funded research. As
22 an example, rural innovation in terrace agriculture developed on the basis of a local coping
23 mechanism and adopted by peasant farmers in Latin America may serve as an adaptation option or
24 starting place for learning about climate change responses (Bocco and Napoletano 2017). Clemens et
25 al. (2015) found that an open dialogue platform enabled horizontal exchange of ideas and alliances for
26 social learning and knowledge-sharing in Vietnam. Improving local technologies in a participatory
27 manner, through on-farm experimentation, farmer-to-farmer exchange, consideration of women and
28 youths, is also relevant in mobilising knowledge and technologies.

29

30 **5.7.4.2 Citizen science**

31 Citizen science has been tested as a useful tool with potential for biodiversity conservation (Schmitz
32 et al. 2015) and mobilising knowledge from society. In food systems, knowledge-holders (e.g.,
33 farmers and pastoralists) are trained to gather scientific data in order to promote conservation and
34 resource management (Fulton et al. 2019) or to conserve and use traditional knowledge in developed
35 countries relevant to climate change adaptation and mitigation through the use of ICT (Calvet-Mir et
36 al. 2018).

37

38 **5.7.4.3 Capacity building and education**

39 Mobilising knowledge may also require significant efforts on capacity building and education to scale
40 up food system responses to climate change. This may involve increasing the capacity of farmers to
41 manage current climate risks and to mitigate and adapt in their local contexts, and of citizens and
42 consumers to understand the links between food demand and climate change emissions and impacts,
43 as well as policy makers to take a systemic view of the issues. Capacity building may also require
44 institutional change. For example, alignment of policies towards sustainable and healthy food systems
45 may require building institutional capacity across policy silos.

1 As a tool for societal transformation, education is a powerful strategy to accelerate changes in the way
2 we produce and consume food. Education refers to early learning and life-long acquisition of skills for
3 higher awareness and actions for solving food system challenges (FAO 2005). Education also entails,
4 vocational training, research and institutional strengthening (Hollinger 2015). Educational focus
5 changes according to the supply side (e.g., crop selection, input resource management, yield
6 improvement, and diversification) and the demand side (nutrition and dietary health implications).
7 Education on food loss and waste spans both the supply and demand sides.

8 In developing countries, extension learning such as Farmer Field Schools – also known as Rural
9 Resources Centers – are established to promote experiential learning on improved production and
10 food transformation (FAO 2016c). While in developed countries, mass education campaigns are rising
11 to reduce food waste, improved diets or acceptable food, and ultimately changes the structure of food
12 industries that is based on the large-scale food products (Heller 2019; UNCCD 2017).

13 The design of new education modules from primary to secondary to tertiary education could help
14 create new jobs in the realm of sustainability (e.g., certification programs). For example, one area
15 could be educating managers of recycling programs for food-efficient cities where food and organic
16 waste are recycled to fertilisers (Jara-Samaniego et al. 2017). Research and education need to be
17 coordinated so that knowledge gaps can be filled and greater trust established in shifting behavior of
18 individuals from conventional options to more sustainable ones. Education campaigns can also
19 influence policy and legislation, and help to advance successful outcomes for climate change
20 mitigation and adaptation in regard to supply-side innovations, technologies, trade, and investment,
21 and demand-side evolution of food choices for health and sustainability, and greater gender equality
22 throughout the entire food system. (Heller 2019).

24 **5.7.5 Knowledge gaps and key research areas**

25 Knowledge gaps around options and solutions and their (co-)benefits and trade-offs are increasingly
26 important now that implementation of mitigation and adaptation measures is scaling up.

27 Research is needed on how a changing climate and interventions to respond to it will affect all aspects
28 of food security, including access, utilisation and stability, not just availability. Knowledge gaps
29 across all the food security pillars are one of the barriers hindering mitigation and adaptation to
30 climate change in the food system and its capacity to deliver food security. The key areas for climate
31 change, food systems, and food security research are enlisted below.

33 **5.7.5.1 Impacts and adaptation**

34 *Climate Services (Food availability).* Agriculture and food security is a priority area for the Global
35 Framework for Climate Services (GFCS) a program of the World Meteorological Organization
36 (WMO). The GFCS enables vulnerable sectors and populations to better manage climate variability
37 and adapt to climate change (Hansen et al. 2018). Global precipitation datasets and remote sensing
38 technologies can be used to detect local to regional anomalies in precipitation as a tool for devising
39 early-warning systems for drought-related impacts, such as famine (Huntington et al. 2017). Crop
40 model improvements are needed for evapotranspiration to guide crop water management in future
41 climate (Cammarano et al. 2016).

42 *Crop and livestock genetics (Food availability, utilisation).* Advances in plant breeding are crucial for
43 enhancing food security under changing climate for a wide variety of crops including fruits and
44 vegetables as well as staples. Genetics improvement is needed in order to breed crops and livestock
45 that can both reduce greenhouse gas emissions, increase drought and heat tolerance (e.g., rice), and
46 enhance nutrition and food security (Nankishore and Farrell 2016; Kole et al. 2015). Many of these

1 characteristics already exist in traditional varieties, including orphan crops and autochthonous breeds,
2 so research is needed to recuperate such varieties and evaluate their potential for adaptation and
3 mitigation.

4 Phenomics-assisted breeding appears to be a promising tool for deciphering the stress responsiveness
5 of crop and animal species (Papageorgiou 2017; Kole et al. 2015; Lopes et al. 2015; Boettcher et al.
6 2015). Initially discovered in bacteria and archaea, CRISPR–Cas9 is an adaptive immune system
7 found in prokaryotes and since 2013 has been used as a genome editing tool in plants. The main use
8 of CRISPR systems is to achieve improved yield performance, biofortification, biotic and abiotic
9 stress tolerance, with rice (*Oryza sativa*) being the most studied crop (Gao 2018; Ricoch et al. 2017).

10 *Climate impact models (Food availability)*. Understanding the full range of climate impacts on staple
11 crops (especially those important in developing countries), fruits and vegetables is missing in the
12 current climate impact models. Further, CO₂ effects on nutrition quality of different crops are just
13 beginning to be parameterised in the models (Müller et al. 2014). Bridging these gap is essential for
14 projecting future dietary diversity, healthy diets, and food security (Bisbis et al. 2018). Crop model
15 improvements are needed for evapotranspiration to guide crop water management in future climate
16 (Cammarano et al. 2016). Similarly, more studies are needed to understand the impacts of climate
17 change on global rangelands and livestock and aquaculture, which have received comparatively less
18 attention than the impacts on crop production.

19 *Resilience to extreme events (Food availability, access, utilisation, and stability)*. On the adaptation
20 side, knowledge gaps include impacts of climate shocks (Rodríguez Osuna et al. 2014) as opposed to
21 impacts of slow-onset climate change, how climate-related harvest failures in one continent may
22 influence food security outcomes in others, impacts of climate change on fruits and vegetables and
23 their nutrient contents.

24

25 **5.7.5.2 Emissions and mitigation**

26 *GHG emissions inventory techniques (Food utilisation)*. Knowledge gaps include food consumption-
27 based emissions at national scales, embedded emissions (overseas footprints) of food systems,
28 comparison of GHG emissions per type of food systems (e.g., smallholder and large-scale commercial
29 food system), and GHG emissions from land-based aquaculture. An additional knowledge gap is the
30 need for more socio-economic assessments of the potential of various integrated practices to deliver
31 the mitigation potential estimated from a biophysical perspective. While studies often project how
32 much CO₂ could theoretically be sequestered in soil, for instance, there is not yet discussion of the
33 potential for this to be effectively monitored, verified, and implemented, once barriers and incentives
34 to adoption of the techniques, practices, and technologies are considered. Thus, future research needs
35 fill the gaps on evaluation of climate actions in the food system.

36 *Food supply chains (Food availability)*. The expansion of the cold chain into developing economies
37 means increased energy consumption and GHG emissions at the consumer stages of the food system,
38 but its net impact on GHG emissions for food systems as a whole is complex and uncertain (Heard
39 and Miller 2016). Further understanding of negative side effects in intensive food processing systems
40 is still needed.

41 Blockchains, as a distributed digital ledger technology which ensures transparency, traceability, and
42 security, is showing promise for easing some global food supply chain management challenges,
43 including the need for documentation of sustainability and the circular economy for stakeholders
44 including governments, communities, and consumers to meet sustainability goals. Blockchain-led
45 transformation of food supply chains is still in its early stages; research is needed on overcoming
46 barriers to adoption (Tripoli and Schmidhuber 2018; Casado-Vara et al. 2018; Mao et al. 2018; Saberi
47 et al. 2019).

1

2 **5.7.5.3 Synergies and trade-offs**

3 *Supply-side and demand-side mitigation and adaptation (Food availability, utilisation).* Knowledge
4 gaps exist on potential and risk associated with novel mitigation technologies on supply side (e.g.,
5 inhibitors, targeted breeding, cellular agriculture, etc.). Additionally, most integrated assessment
6 models (IAMs) currently have limited regional data on BECCS projects because of little BECCS
7 implementation (Lenzi et al. 2018). Hence, several BECCS scenarios seem to rely on unrealistic
8 assumptions regarding regional climate, soils and infrastructure suitability (Köberle et al. 2019) as
9 well as trade of international trade of bioenergy (Lamers et al. 2011).

10 Areas for study include how to incentivise, regulate, and raise awareness on the co-benefits of healthy
11 consumption patterns and climate change mitigation and adaptation; to improve access to healthy
12 diets for vulnerable groups through food assistance programs; and to implement policies and
13 campaigns to reduce food loss and food waste. Knowledge gaps also exist on the role of different
14 policies, and underlying uncertainties, to promote changes in food habits towards climate resilience
15 and healthy diets.

16 *Food systems, land use change, and telecoupling (Food availability, access, utilisation).* The
17 analytical framework of telecoupling has recently been proposed to address this complexity,
18 particularly the connections, flows, and feedbacks characterising food systems (Friis et al. 2016;
19 Easter et al. 2018). For example, how will climate-induced shifts in livestock and crop diseases affect
20 food production and consumption in the future. Investigating the social and ecological consequences
21 of these changes will contribute to decision making under uncertainty in the future. Research areas
22 include food systems and their boundaries, hierarchies, and scales through metabolism studies,
23 political ecology and cultural anthropology.

24 *Food-Energy-Water Nexus (Food availability, utilisation, stability).* Emerging interdisciplinary
25 science efforts are providing new understanding of the interdependence of food, energy, and water
26 systems and these interdependencies are beginning to take into account climate change, food security,
27 and AFOLU assessments (Scanlon et al. 2017; Liu et al. 2017). These science advances, in turn,
28 provide critical information for coordinated management to improve the affordability, reliability, and
29 environmental sustainability of food, energy, and water systems. Despite significant advances within
30 the past decade, there are still many challenges for the scientific community. These include the need
31 for interdisciplinary science related to the food-energy-water nexus; ground-based monitoring and
32 modelling at local-to-regional scales (Van Gaalen et al. 2017); incorporating human and institutional
33 behaviour in models; partnerships among universities, industry, and government to develop policy-
34 relevant data; and systems modelling to evaluate trade-offs associated with food-energy-water
35 decisions (Scanlon et al. 2017). However, the nexus approach, as a conceptual framework, requires
36 the recognition that, although land and the goods and services it provides is finite, potential demand
37 for the goods and services may be greater than the ability to supply them sustainably (Benton et al.
38 2018). By addressing demand-side issues, as well as supply-side efficiencies, it provides a potential
39 route for minimising trade-offs for different goods and services (Benton et al. 2018) and (Section 5.6).

40

41 **5.8 Future challenges to food security**

42 A particular concern in regard to the future of food security is the potential for the impacts of
43 increasing climate extremes on food production to contribute to multi-factored complex events such
44 as food price spikes. In this section, we assess literature on food price spikes and potential strategies
45 for increasing resilience to such occurrences. We then assess the potential for such food system events
46 to affect migration and conflict.

1

2 **5.8.1 Food price spikes**

3 Under average conditions, global food system markets may function well, and equilibrium approaches
4 can estimate demand and supply with some confidence; however, if there is a significant shock, the
5 market can fail to smoothly link demand and supply through price, and a range of factors can act to
6 amplify the effects of the shock, and transmit it across the world (Box 5.5). Given the potential for
7 shocks driven by changing patterns of extreme weather to increase with climate change, there is the
8 potential for market volatility to disrupt food supply through creating food price spikes. This potential
9 is exacerbated by the interconnectedness of the food system (Puma et al. 2015) with other sectors (i.e.,
10 the food system depends on water, energy, transport, etc.) (Homer-Dixon et al. 2015), so the impact of
11 shocks can propagate across sectors and geographies (Homer-Dixon et al. 2015). There is also less
12 spare land globally than there has been in the past, such that if prices spike, there are fewer options to
13 bring new production on stream (Marianela et al. 2016).

14 Increasing extreme weather events can disrupt production and transport logistics. For example, in
15 2012 the US Corn Belt suffered a widespread drought; US corn yield declined 16% compared to 2011
16 and 25% compared to 2009. A record yield loss of 2016 in French that is attributed to a conjunction of
17 abnormal warmth in late autumn and abnormal wet in the following spring (Ben-Ari et al. 2018) is
18 another well-documented example. To the extent that such supply shocks are associated with climate
19 change, they may become more frequent and contribute to greater instability in agricultural markets in
20 the future. Furthermore, analogue conditions of past extremes might create significantly greater
21 impacts in a warmer world. A study simulating analogous conditions to the Dustbowl drought in
22 today's agriculture suggests that Dust-Bowl-type droughts today would have unprecedented
23 consequences, with yield losses about 50% larger than the severe drought of 2012 (Glotter and Elliott
24 2016). Damages at these extremes are highly sensitive to temperature, worsening by about 25% with
25 each degree centigrade of warming. By mid-century, over 80% of summers are projected to have
26 average temperatures that are likely to exceed the hottest summer in the Dustbowl years (1936)
27 (Glotter and Elliott 2016).

28 How a shortfall in production – or an interruption in trade due to an event affecting a logistics choke-
29 point (Wellesley et al. 2017) – of any given magnitude may create impacts depends on many
30 interacting factors (Homer-Dixon et al. 2015; Tadasse et al. 2016; Challinor et al. 2018). The principal
31 route is by affecting agricultural commodity markets, which respond to a perturbation through
32 multiple routes as in Figure 5.17 **Error! Reference source not found.** This includes pressures from
33 other sectors (such as if biofuels policy is incentivising crops for the production of ethanol, as
34 happened in 2007–2008). The market response can be amplified by poor policies, setting up trade and
35 non-trade barriers to exports, from countries seeking to ensure their local food security (Bailey et al.
36 2015). Furthermore, the perception of problems can fuel panic buying on the markets that in turn
37 drives up prices.

38 Thus, the impact of an extreme weather event on markets has both a *trigger* component (the event)
39 and a *risk perception* component (Challinor et al. 2016, 2018). Through commodity markets, prices
40 change across the world because almost every country depends, to a greater or lesser extent, on trade
41 to fulfil local needs. Commodity prices can also affect local market prices by altering input prices,
42 changing the cost of food aid, and through spill-over effects; for example, in 2007–2008 the grain
43 affected by extreme weather was wheat, but there was a significant price spike in rice markets (Dawe
44 2010).

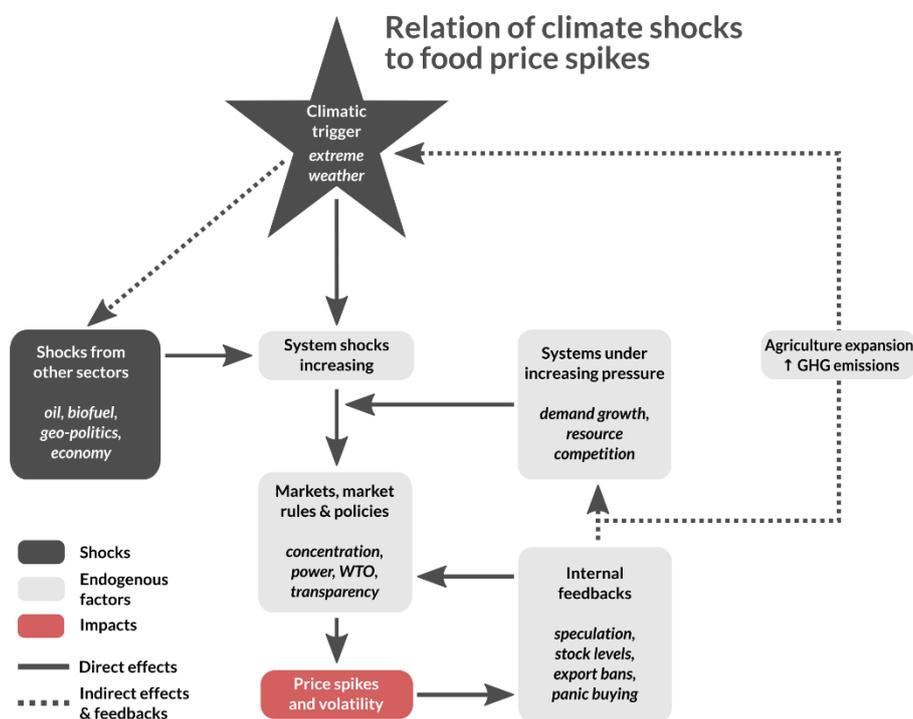
45 As discussed by Bailey et al. (2015), there are a range of adaptation measures that can be put in place
46 to reduce the impact of climate-related production shortfalls. These include (a) ensuring transparency
47 of public and private stocks, as well as improved seasonal forecasting to signal forthcoming yield

1 shortfalls (FAO 2016a; Ceglar et al. 2018; Iizumi et al. 2018), (b) building real or virtual
 2 stockholdings, (c) increasing local productivity and diversity (as a hedge against a reliance on trade)
 3 and (d) ensuring smoother market responses, through, for example, avoiding the imposition of export
 4 bans.

5 In summary, given the likelihood that extreme weather will increase, in both frequency and magnitude
 6 (Hansen et al. 2012; Coumou et al. 2014; Mann et al. 2017; Bailey et al. 2015), and the current state
 7 of global and cross-sectoral interconnectedness, the food system is at increasing risk of disruption
 8 (*medium evidence, medium agreement*), with large uncertainty about how this could manifest. There
 9 is therefore a need to build resilience into international trade as well as local supplies.

10

11



12

13 **Figure 5.17 Underlying processes that affect the development of a food price spike in agricultural**
 14 **commodity markets (Challinor et al. 2018)**

15 **Box 5.5 Market drivers and the consequences of extreme weather in 2010-2011**

16 The 2010–2011 food price spike was initially triggered by the exceptional heat in summer 2010, with
 17 an extent from Europe to the Ukraine and Western Russia (Barriopedro et al. 2011; Watanabe et al.
 18 2013; Hoag 2014). The heatwave in Russia was extreme in both temperature (over 40°C) and duration
 19 (from July to mid-August in 2010). This reduced wheat yields by approximately one third (Wegren
 20 2011; Marchand et al. 2016). Simultaneously, in the Indus Valley in Pakistan, unprecedented rainfall
 21 led to flooding, affecting the lives and livelihoods of 20 million people. There is evidence that these
 22 effects were both linked and made more likely through climate change (Mann et al. 2017).

23 In response to its shortfall in yields, Russia imposed an export ban in order to maintain local food
 24 supplies. Other countries responded in a largely uncoordinated ways, each of them driven by
 25 internal politics as well as national self-interests (Jones and Hiller 2017). Overall, these measures
 26 led to rapid price rises on the global markets (Welton 2011), partly through panic buying, but also
 27 through financial speculation (Spratt 2013).

1 Analysis of responses to higher food prices in the developing world showed that lower-income groups
2 responded by taking on more employment, reducing food intake, limiting expenditures, spending
3 savings (if available), and participating in demonstrations. People often identified their problems as
4 stemming from collusion between powerful incumbent interests (e.g., of politicians and big business)
5 and disregard for the poor (Hossain and Green 2011). This politicised social response helped spark
6 food-related civil protest, including riots, across a range of countries in 2010–2011 (Natalini et al.
7 2017). In Pakistan, food price rises were exacerbated by the economic impacts of the floods, and
8 which further contributed to food-related riots in 2010.

9 Price spikes also impact on food security in the developed world. In the UK, global commodity price
10 inflation influenced local food prices, increasing food-price inflation by ~5 times at the end of 2010.
11 Comparing household purchases over the five year period from 2007 to 2011 showed that the amount
12 of food bought declined, on average, by 4.2%, whilst paying 12% more for it. The lowest income
13 decile spent 17% more by 2011 than they did in 2007 (Holding et al. 2013; Tadasse et al. 2016).
14 Consumers also saved money by trading down for cheaper alternatives. For the poorest, in the
15 extreme situation, food became unaffordable: the Trussell Trust, a charity supplying emergency food
16 handouts for people in crisis, noted a 50% increase in handouts in 2010.

18 **5.8.2 Migration and conflict**

19 Since the IPCC AR5 (Porter et al. 2014; Cramer et al. 2014), new work has advanced multi-factor
20 methodological issues related to migration and conflict (e.g., Kelley et al. 2015, 2017; Werrell et al.
21 2015; Challinor et al. 2018; Pasini et al. 2018). These in particular have addressed systemic risks to
22 food security that result from cascading impacts triggered by droughts and floods and how these are
23 related to a broad range of societal influences.

24 Climate variability and extremes have short-, medium- and long-term impacts on livelihoods and
25 livelihood assets – especially of the poor – contributing to greater risk of food insecurity and
26 malnutrition (FAO et al. 2018). Drought threatens local food security and nutrition and aggravates
27 humanitarian conditions, which can trigger large-scale human displacement and create a breeding
28 ground for conflict (Maystadt and Ecker 2014). There is *medium agreement* that existing patterns of
29 conflict could be reinforced under climate change affecting food security and livelihood opportunities,
30 for example, already fragile regions with ethnic divides such as North and Central Africa as well as
31 Central Asia (Buhaug 2016; Schleussner et al. 2016) (Box 5.6).

32 Challinor et al. (2018) have developed a typology for transboundary and transboundary risk
33 transmission that distinguishes the roles of climate and social and economic systems. To understand
34 these complex interactions, they recommend a combination of methods that include expert judgement;
35 interactive scenario building; global systems science and big data; and innovative use of climate and
36 integrated assessment models; and social science techniques (e.g., surveys, interviews, and focus
37 groups).

39 **5.8.2.1 Migration**

40 There has been a surge in international migration in recent years, with around five million people
41 migrating permanently in 2016 (OECD 2017). Though the initial driver of migration may differ across
42 populations, countries and contexts, migrants tend to seek the same fundamental objective: to provide
43 security and adequate living conditions for their families and themselves. Food insecurity is a critical
44 ‘push’ factor driving international migration, along with conflict, income inequality, and population

1 growth. The act of migration itself causes food insecurity, given the lack of income opportunities and
2 adverse conditions compounded by conflict situations.

3 Warner et al. (2012) found the interrelationships between changing rainfall patterns, food and
4 livelihood security in eight countries in Asia, Africa and Latin America. Several studies in Africa
5 have found that persistent droughts and land degradation contributed to both seasonal and permanent
6 migration (Gray 2011; Gray and Mueller 2012; Hummel 2015; Henry et al. 2004; Folami and Folami
7 2013), worsening contextual vulnerability conditions of different households (Dasgupta et al. 2014).

8 Dependency on rainfed agriculture is from 13% in Mexico to more than 30% in Guatemala,
9 Honduras, and Nicaragua, suggesting a high degree of sensitivity to climate variability and change,
10 and undermined food security (Warner et al. 2009). Studies have demonstrated that Mexican
11 migration (Feng et al. 2010; Nawrotzki et al. 2013) and Central American migration (WFP 2017)
12 fluctuate in response to climate variability. The food system is heavily dependent on maize and bean
13 production and long-term climate change and variability significantly affect the productivity of these
14 crops and the livelihoods of smallholder farmers (WFP 2017). In rural Ecuador, adverse
15 environmental conditions prompt out-migration, although households respond to these challenges in
16 diverse ways resulting in complex migratory responses (Gray and Bilsborrow 2013).

17 Migration patterns have been linked to heat stress in Pakistan (Mueller et al. 2014) and climate
18 variability in the Sundarbans due to decline in food security (Guha and Roy 2016). In Bangladesh, the
19 impacts of climate change have been on the rise throughout the last three decades with increasing
20 migration, mostly of men leaving women and children to cope with increasing effects of natural
21 disasters (Rabbani et al. 2015).

22

23 **Box 5.6 Migration in the Pacific region: Impacts of climate change on food security**

24 Climate change-induced displacement and migration in the Pacific has received wide attention in the
25 scientific discourse (Fröhlich and Klepp 2019). The processes of climate change and their effects in
26 the region have serious implications for Pacific Island nations as they influence the environments that
27 are their 'life-support systems' (Campbell 2014). Climate variability poses significant threats to both
28 agricultural production and food security. Rising temperatures and reductions in groundwater
29 availability, as well as increasing frequency and severity of disaster events translate into substantial
30 impacts on food security causing human displacement, a trend that will be aggravated by future
31 climate impacts (ADB 2017). Declining soil productivity, groundwater depletion, and non-availability
32 of freshwater threatens agricultural production in many remote atolls.

33 Many countries in the Pacific devote a large share of available land area to agricultural production.
34 For example, more than 60% of land area is cultivated in the Marshall Islands and Tuvalu and more
35 than 40% in Kiribati and Tonga. With few options to expand agricultural area, the projected impacts
36 of climate change on food production are of particular concern (ADB 2013, 2017). The degradation of
37 available land area for traditional agriculture, adverse disruptions of agricultural productivity and
38 diminishing livelihood opportunities through climate change impacts leads to increasing poverty and
39 food insecurity, incentivising migration to urban agglomerations (ADB 2017; FAO et al. 2018).

40 Campbell (2014) describe the trends that lead to migration. First, climate change, including sea level
41 rise, affect communities' land security, which is the physical presence on which to live and sustain
42 livelihoods. Second, they impinge on livelihood security (especially food security) of island
43 communities where the productivity of both subsistence and commercial food production systems is
44 reduced. Third, the effects of climate change are especially severe on small-island environments since
45 they result in declining ecological habitat. The effects on island systems are mostly manifested in
46 atolls through erosion and inundation, and on human populations through migration. Population

1 growth and scenarios of climate change is *likely* to further induce food stress as impacts unfolds in
2 coming decades (Campbell 2015).

3 While the populations of several islands and island groups in the Pacific (e.g., Tuvalu, Carteret
4 Islands, and Kiribati) have been perceived as the first probable victims of rising seas so that their
5 inhabitants would become, and in some quarters already are seen to be, the first ‘environmental’ or
6 ‘climate change refugees,’ migration patterns vary. Especially in small islands, the range and nature of
7 the interactions among economic, social, and/or political drivers are complex. For example, in the
8 Maldives, Stojanov et al. (2017) show that while collective perceptions support climate change
9 impacts as being one of the key factors prompting migration, individual perceptions give more
10 credence to other cultural, religious, economic or social factors.

11 In the Pacific, Tuvalu has long been a prime candidate to disappear due to rising sea level, forcing
12 human migration. However, results of a recent study (Kench et al. 2018) challenge perceptions of
13 island loss in Tuvalu, reporting despite sea level rise, a net increase in land area of 73.5 ha. The
14 findings suggest that islands are dynamic features likely to persist as habitation sites over the next
15 century, presenting opportunities for adaptation that embrace the heterogeneity of island types and
16 processes. Farbotko (2010) and Farbotko and Lazrus (2012) present Tuvalu as a site of ‘wishful
17 sinking,’ in the climate change discourse. These authors argue that representations of Tuvalu as a
18 laboratory for global climate change migration are visualisations by non-locals.

19 In Nanumea (Tuvalu), forced displacements and voluntary migrations are complex decisions made
20 by individuals, families and communities in response to discourses on risk, deteriorating
21 infrastructure and other economic and social pressures (Marino and Lazrus 2015). In many atoll
22 nations in western Pacific, migration has increasingly become a sustainable livelihood strategy,
23 irrespective of climate change (Connell 2015).

24 In Lamén Bay, Vanuatu, migration is both a cause and consequence of local vulnerabilities. While
25 migration provides an opportunity for households to meet their immediate economic needs, it limits
26 the ability of the community to foster longer-term economic development. At the same time,
27 migration adversely affects the ability of the community to maintain food security due to lost labour
28 and changing attitudes towards traditional ways of life among community members (Craven 2015).

29 Small islands are very sensitive to climate change impacts (*high confidence*) (Nurse et al. 2014) and
30 impacted by multiple climatic stressors (see (IPCC 2018a) and SROCC). Food security in the Pacific,
31 especially in Micronesia, has worsened in the past half century and climate change is *likely* to further
32 hamper local food production, especially in low-lying atolls (Connell 2016) Migration in small islands
33 (internally and internationally) occurs for multiple reasons and purposes, mostly for better livelihood
34 opportunities (*high confidence*).

35 Beyond sea level rise, effects of increasing frequency and intensity of extreme events such as severe
36 tropical cyclones are *likely* to affect human migration in the Pacific (Connell 2015; Krishnapillai and
37 Gavenda 2014; Charan et al. 2017; Krishnapillai 2017). On Yap Island, extreme weather events are
38 affecting every aspect of atoll communities’ existence, mainly due to islands’ small size, their low
39 elevation, and extensive coastal areas (Krishnapillai 2018). Displaced atoll communities on Yap
40 Island grow a variety of nutritious vegetables and use alternative crop production methods such as
41 small-plot intensive farming, raised bed gardening, as part of a community-based adaptation program
42 (Krishnapillai and Gavenda 2014; Krishnapillai 2018).

43 Recurrences of natural disasters and crises threaten food security through impacts on traditional
44 agriculture, causing the forced migration and displacement of coastal communities to highlands in
45 search of better living conditions. Although considerable differences occur in the physical
46 manifestations of severe storms, such climate stressors threaten the life-support systems of many atoll

1 communities (Campbell et al. 2014). Failure of these systems resulting from climate disasters propel
2 vulnerable atoll communities into poverty traps, and low adaptive capacity could eventually force
3 these communities to migrate.

4 5 **5.8.2.2 Conflict**

6 While climate change will not alone cause conflict, it is often acknowledged as having the potential to
7 exacerbate or catalyse conflict in conjunction with other factors. Increased resource competition can
8 aggravate the potential for migration to lead to conflict. When populations continue to increase,
9 competition for resources will also increase, and resources will become even scarcer due to climate
10 change (Hendrix and Glaser 2007). In agriculture-dependent communities in low-income contexts,
11 droughts have been found to increase the likelihood of violence and prolonged conflict at the local
12 level, which eventually pose a threat to societal stability and peace (FAO et al. 2017). In contrast,
13 conflicts can also have diverging effects on agriculture due to land abandonment, resulting in forest
14 growth, or agriculture expansion causing deforestation, e.g., in Colombia (Landholm et al. 2019).

15 Several studies have explored the causal links among climate change, drought, impacts on agricultural
16 production, livelihoods, and civil unrest in Syria from 2007-2010, but without agreement as to the role
17 played by climate in subsequent migration (Kelley et al. 2015, 2017; Challinor et al. 2018; Selby et al.
18 2017; Hendrix 2018). Contributing factors that have been examined include rainfall deficits,
19 population growth, agricultural policies, and influx of refugees that had placed burdens on the
20 region's water resources (Kelley et al. 2015). Drought may have played a role as a trigger, as this
21 drought was the longest and the most intense in the last 900 years (Cook et al. 2016; Mathbout et al.
22 2018). Some studies linked the drought to widespread crop failure, but the climate hypothesis has
23 been contested (Selby et al. 2017; Hendrix 2018). Recent evidence shows that the severe drought
24 triggered agricultural collapse and displacement of rural farm families with approximately 300,000
25 families going to Damascus, Aleppo and other cities (Kelley et al. 2017).

26 Persistent drought in Morocco during the early 1980s resulted in food riots and contributed to an
27 economic collapse (El-Said and Harrigan 2014). A drought in Somalia that fuelled conflict through
28 livestock price changes, establishing livestock markets as the primary channel of impact (Maystadt
29 and Ecker 2014). Cattle raiding as a normal means of restocking during drought in the Great Horn of
30 Africa led to conflict (ICPAC and WFP 2017) whereas a region-wide drought in northern Mali in
31 2012 wiped out thousands of livestock and devastated the livelihoods of pastoralists, in turn swelling
32 the ranks of armed rebel factions and forcing others to steal and loot for survival (Breisinger et al.
33 2015).

34 On the other hand, inter-annual adjustments in international trade can play an important role in
35 shifting supplies from food surplus regions to regions facing food deficits which emerge as a
36 consequence of extreme weather events, civil strife, and/or other disruptions (Baldos and Hertel
37 2015). A more freely functioning global trading system is tested for its ability to deliver improved
38 long run food security in 2050.

39 In summary, given increasing extreme events and global and cross-sectoral interconnectedness, the
40 food system is at increasing risk of disruption, e.g., via migration and conflict (*high confidence*).
41 {5.2.3, 5.2.4}

43 **Frequently Asked Questions**

44 **FAQ 5.1 How does climate change affect food security?**

1 Climate change negatively affects all four pillars of food security: availability, access, utilisation and
2 stability. Food availability may be reduced by negative climate change impacts on productivity of
3 crops, livestock and fish, for instance due to increases in temperature and changes in rainfall patterns.
4 Productivity is also negatively affected by increased pests and diseases, as well as changing
5 distributions of pollinators under climate change. Food access and its stability may be affected
6 through disruption of markets, prices, infrastructure, transport, manufacture, and retail, as well as
7 direct and indirect changes in income and food purchasing power of low-income consumers. Food
8 utilisation may be directly affected by climate change due to increases in mycotoxins in food and feed
9 with rising temperatures and increased frequencies of extreme events, and indirectly through effects
10 on health. Elevated atmospheric CO₂ concentrations can increase yields at lower temperature
11 increases, but tend to decrease protein content in many crops, reducing their nutritional values.
12 Extreme events, e.g., flooding, will affect the stability of food supply directly through disruption of
13 transport and markets.

14 15 **FAQ 5.2 How can changing diets help address climate change?**

16 Agricultural activities emit substantial amounts of greenhouse gases (GHGs). Food supply chains
17 activities past the farm gate (e.g., transportation, storage, packaging) also emit GHGs, for instance due
18 to consumption of energy. GHG emissions from food production vary across food types. Producing
19 animal-sourced food (i.e., meat and dairy) emits larger amount of GHGs than growing crops,
20 especially in intensive, industrial livestock systems. This is mainly true for commodities produced by
21 ruminant livestock such as cattle, due to enteric fermentation processes that are large emitters of
22 methane. Changing diets towards a lower share of animal-sourced food, once implemented at scale,
23 reduces the need to raise livestock and changes crop production from animal feed to human food. This
24 reduces the need for agricultural land compared to present and thus generates changes in the current
25 food system. From field to consumer this would reduce overall GHG emissions. Changes in consumer
26 behaviour beyond dietary changes can also have, at scale, effects on overall GHG emissions from
27 food systems. Consuming regional and seasonal food can reduce GHG emissions, if they are grown
28 efficiently.

29 30 **References**

- 31 Abarca-Gómez, L., and Coauthors, 2017: Worldwide trends in body-mass index, underweight,
32 overweight, and obesity from 1975 to 2016: a pooled analysis of 2416 population-based
33 measurement studies in 128.9 million children, adolescents, and adults. *Lancet*,
34 doi:10.1016/S0140-6736(17)32129-3.
35 <http://linkinghub.elsevier.com/retrieve/pii/S0140673617321293> (Accessed October 26, 2017).
- 36 Abbas, G., and Coauthors, 2017: Quantification the impacts of climate change and crop management
37 on phenology of maize-based cropping system in Punjab, Pakistan. *Agric. For. Meteorol.*, **247**,
38 42–55, doi:10.1016/j.agrformet.2017.07.012.
39 <https://linkinghub.elsevier.com/retrieve/pii/S0168192317302289> (Accessed November 4, 2018).
- 40 Aberman, N.-L., and C. Tirado, 2014: Impacts of Climate Change on Food Utilization. *Glob. Environ.*
41 *Chang.*, 717–724, doi:10.1007/978-94-007-5784-4. <http://link.springer.com/10.1007/978-94-007-5784-4>.
- 43 Abiona, B. G., E. O. Fakoya, and J. Esun, 2016: The Impacts of Climate Change on the Livelihood of
44 Arable Crop Farmers in Southwest, Nigeria. Springer, Cham, 289–296
45 http://link.springer.com/10.1007/978-3-319-25814-0_20 (Accessed November 4, 2018).
- 46 ADB, 2013: Food Security Challenges in Asia. 2014.

- 1 —, 2014: *TECHNOLOGIES TO SUPPORT CLIMATE CHANGE ADAPTATION IN*
2 *DEVELOPING ASIA.*
3 <https://www.google.com/url?sa=t&rct=j&q=&esrc=s&source=web&cd=1&cad=rja&uact=8&ved=2ahUKEwi7zIaYvaveAhWHF8AKHcjUA1wQFjAAegQICRAB&url=https%3A%2F%2Fwww.adb.org%2Fpublications%2Ftechnologies-support-climate-change-adaptation-developing-asia&usg=AOvVaw3eRGXZ>.
- 7 —, 2017: *Food Insecurity in Asia: Why Institutions Matter.* 415 pp pp.
8 [https://www.adb.org/sites/default/files/publication/366791/adbi-food-insecurity-asia-why-](https://www.adb.org/sites/default/files/publication/366791/adbi-food-insecurity-asia-why-institutions-matter.pdf)
9 [institutions-matter.pdf](https://www.adb.org/sites/default/files/publication/366791/adbi-food-insecurity-asia-why-institutions-matter.pdf).
- 10 Adger, W. N., J. Barnett, K. Brown, N. Marshall, and K. O'Brien, 2013: Cultural dimensions of
11 climate change impacts and adaptation. *Nat. Clim. Chang.*, **3**, 112–117,
12 doi:10.1038/nclimate1666. <http://www.nature.com/articles/nclimate1666> (Accessed January 11,
13 2018).
- 14 Adhikari, L., A. Hussain, and G. Rasul, 2017: Tapping the potential of neglected and underutilized
15 food crops for sustainable nutrition security in the mountains of Pakistan and Nepal. *Sustain.*, **9**,
16 doi:10.3390/su9020291.
- 17 Adiku, S. G. K., and Coauthors, 2015: Climate change impacts on west african agriculture: an
18 integrated regional assessment (CIWARA). *Handbook of Climate Change and agroecosystems:*
19 *The Agricultural Model Intercomparison and Improvement Project Integrated Crop and*
20 *Economic Assessments, Part 2*, 25–73.
- 21 Agarwal, C., G. M. Green, J. M. Grove, T. P. Evans, and C. M. Schweik, 2002: *A review and*
22 *assessment of land-use change models: dynamics of space, time, and human choice.* Newton
23 Square, PA: U.S., 61 pp. <https://www.fs.usda.gov/treesearch/pubs/5027> (Accessed June 5,
24 2018).
- 25 Aggarwal, P. K., and Coauthors, 2018: The climate-smart village approach: framework of an
26 integrative strategy for scaling up adaptation options in agriculture. *Ecol. Soc.*, **23**, 14.
- 27 Aghakouchak, A., A. Farahmand, F. S. Melton, J. Teixeira, M. C. Anderson, B. D. Wardlow, and C.
28 R. Hain, 2015: Reviews of Geophysics Remote sensing of drought: Progress, challenges. *Rev.*
29 *Geophys*, **53**, 452–480, doi:10.1002/2014RG000456. Received.
- 30 AGRA, 2017: *Africa Agriculture Status Report: The Business of Smallholder Agriculture in Sub-*
31 *Saharan Africa.* Nairobi, Kenya, 180 pp. [https://agra.org/wp-content/uploads/2017/09/Final-](https://agra.org/wp-content/uploads/2017/09/Final-AASR-2017-Aug-28.pdf)
32 [AASR-2017-Aug-28.pdf](https://agra.org/wp-content/uploads/2017/09/Final-AASR-2017-Aug-28.pdf).
- 33 Agrawala, S., and M. Carraro, 2010: Assessing the Role of Microfinance in Fostering Adaptation to
34 Climate Change. *Ssrn*, doi:10.2139/ssrn.1646883.
- 35 Ahlgren, I., S. Yamada, and A. Wong, 2014: Rising Oceans, Climate Change, Food Aid and Human
36 Rights in the Marshall Islands. *Heal. Hum. Rights J.*, **16**, 69–81,
37 doi:10.2307/healhumarigh.16.1.69. <https://www.jstor.org/stable/pdf/healhumarigh.16.1.69.pdf>
38 (Accessed October 30, 2018).
- 39 Ahmad, A., and Coauthors, 2015: Impact of climate change on the rice--wheat cropping system of
40 Pakistan. *Handbook of Climate Change and Agroecosystems: The Agricultural Model*
41 *Intercomparison and Improvement Project Integrated Crop and Economic Assessments, Part 2*,
42 World Scientific, 219–258.
- 43 Ahmed, N., A. Occhipinti-Ambrogi, and J. F. Muir, 2013: The impact of climate change on prawn
44 postlarvae fishing in coastal Bangladesh: Socioeconomic and ecological perspectives. *Mar.*
45 *Policy*, **39**, 224–233, doi:10.1016/j.marpol.2012.10.008.
- 46 —, S. W. Bunting, M. Glaser, M. S. Flaherty, and J. S. Diana, 2017: Can greening of aquaculture
47 sequester blue carbon? *Ambio*, **46**, 468–477, doi:10.1007/s13280-016-0849-7.
48 <http://link.springer.com/10.1007/s13280-016-0849-7> (Accessed April 14, 2019).

- 1 —, S. Thompson, and M. Glaser, 2018: Integrated mangrove-shrimp cultivation: Potential for blue
2 carbon sequestration. *Ambio*, **47**, 441–452, doi:10.1007/s13280-017-0946-2.
3 <http://www.ncbi.nlm.nih.gov/pubmed/28980188> (Accessed April 14, 2019).
- 4 Aipira, C., A. Kidd, and K. Morioka, 2017: Climate Change Adaptation in Pacific Countries:
5 Fostering Resilience Through Gender Equality. Springer, Cham, 225–239
6 http://link.springer.com/10.1007/978-3-319-50094-2_13 (Accessed April 7, 2019).
- 7 Ajani, E., E. Onwubuya, and R. Mgbenka, 2013: Approaches to Economic Empowerment of Rural
8 Women for Climate Change Mitigation and Adaptation: Implications for Policy. *J. Agric. Ext.*,
9 **17**, 23, doi:10.4314/jae.v17i1.3. <http://www.ajol.info/index.php/jae/article/view/90252>
10 (Accessed March 13, 2019).
- 11 Akram-Lodhi, A. H., and others, 2013: *Hungry for Change: farmers, food justice and the agrarian*
12 *question*. Kumarian Press Inc.,.
- 13 Albrecht, T. R., A. Crootof, and C. A. Scott, 2018: The Water-Energy-Food Nexus: A systematic
14 review of methods for nexus assessment. *Environ. Res. Lett.*, **13**, 43002, doi:10.1088/1748-
15 9326/aaa9c6. <http://dx.doi.org/10.1088/1748-9326/aaa9c6>.
- 16 Alexander, P., C. Brown, A. Arneth, J. Finnigan, and M. D. A. Rounsevell, 2016: Human
17 appropriation of land for food: The role of diet. *Glob. Environ. Chang.*, **41**, 88–98,
18 doi:10.1016/J.GLOENVCHA.2016.09.005.
19 <https://www.sciencedirect.com/science/article/abs/pii/S0959378016302370?via%3Dihub>
20 (Accessed March 9, 2019).
- 21 —, —, —, —, D. Moran, and M. D. A. A. Rounsevell, 2017: Losses, inefficiencies and
22 waste in the global food system. *Agric. Syst.*, **153**, 190–200, doi:10.1016/j.agry.2017.01.014.
23 <http://dx.doi.org/10.1016/j.agry.2017.01.014>.
- 24 Ali, A., W. K. Yeoh, C. Forney, and M. W. Siddiqui, 2018: Advances in postharvest technologies to
25 extend the storage life of minimally processed fruits and vegetables. *Crit. Rev. Food Sci. Nutr.*,
26 **58**, 2632–2649, doi:10.1080/10408398.2017.1339180.
27 <http://dx.doi.org/10.1080/10408398.2017.1339180>.
- 28 Allen, A., L. Griffin, and C. Johnson, 2017: *Environmental justice and urban resilience in the global*
29 *south*. 1–307 pp.
- 30 Alston, M., 2009: Drought policy in Australia: gender mainstreaming or gender blindness? *Gender,*
31 *Place Cult.*, **16**, 139–154, doi:10.1080/09663690902795738.
32 <http://www.tandfonline.com/doi/full/10.1080/09663690902795738> (Accessed November 3,
33 2018).
- 34 —, 2014: Gender mainstreaming and climate change. *Womens. Stud. Int. Forum*, **47**, 287–294,
35 doi:10.1016/J.WSIF.2013.01.016.
36 <https://www.sciencedirect.com/science/article/abs/pii/S0277539513000204> (Accessed October
37 29, 2018).
- 38 Altieri, M., C. Nicholls, R. Montalba, M. A. Altieri, C. I. Nicholls, and R. Montalba, 2017:
39 Technological Approaches to Sustainable Agriculture at a Crossroads: An Agroecological
40 Perspective. *Sustainability*, **9**, 349, doi:10.3390/su9030349. [http://www.mdpi.com/2071-
41 1050/9/3/349](http://www.mdpi.com/2071-1050/9/3/349) (Accessed March 3, 2019).
- 42 Altieri, M. A., 2002: Agroecology: the science of natural resource management for poor farmers in
43 marginal environments. *Agric. Ecosyst. Environ.*, **93**, 1–24, doi:10.1016/S0167-8809(02)00085-
44 3. <https://www.sciencedirect.com/science/article/pii/S0167880902000853> (Accessed October
45 27, 2018).
- 46 Altieri, M. A., and P. Koohafkan, 2008: *Enduring Farms: Climate Change, Smallholders and*
47 *Traditional Farming Communities*. Third World Network, Pinang, 72 pp.
- 48 —, and C. I. Nicholls, 2017: The adaptation and mitigation potential of traditional agriculture in a

- 1 changing climate. *Clim. Change*, **140**, 33–45, doi:10.1007/s10584-013-0909-y.
2 <http://link.springer.com/10.1007/s10584-013-0909-y> (Accessed May 18, 2018).
- 3 —, P. Rosset, and L. A. Thrupp, 1998: *The potential of agroecology to combat hunger in the*
4 *developing world*. Washington D.C., 1–2 pp.
5 <http://ebrary.ifpri.org/utils/getfile/collection/p15738coll2/id/126481/filename/126692.pdf>.
- 6 —, C. I. Nicholls, A. Henao, and M. A. Lana, 2015: Agroecology and the design of climate change-
7 resilient farming systems. *Agron. Sustain. Dev.*, **35**, 869–890, doi:10.1007/s13593-015-0285-2.
- 8 Amanullah, A., and S. Khalid, 2016: Integrated Use of Phosphorus, Animal Manures and
9 Biofertilizers Improve Maize Productivity under Semiarid Condition. *Organic Fertilizers*, p. 20
10 pp.
- 11 American Diabetes Association, 2018: Economic Costs of Diabetes in the US in 2017. *Diabetes Care*,
12 **41**, 917.
- 13 Anderson, K., and G. Peters, 2016: The trouble with negative emissions. *Science* (80-.), **354**, 182–
14 183, doi:10.1126/science.aah4567. <http://www.ncbi.nlm.nih.gov/pubmed/27738161> (Accessed
15 November 4, 2018).
- 16 Antle, J. M., and Coauthors, 2015: AgMIP’s transdisciplinary agricultural systems approach to
17 regional integrated assessment of climate impacts, vulnerability, and adaptation. *Handbook of*
18 *Climate Change and Agroecosystems: The Agricultural Model Intercomparison and*
19 *Improvement Project Integrated Crop and Economic Assessments, Part 1*, World Scientific, 27–
20 44.
- 21 Antwi-Agyei, P., A. J. Dougill, and L. C. Stringer, 2015: Impacts of land tenure arrangements on the
22 adaptive capacity of marginalized groups: The case of Ghana’s Ejura Sekyedumase and Bongo
23 districts. *Land use policy*, doi:10.1016/j.landusepol.2015.08.007.
- 24 Aqeel-Ur-Rehman, A. Z. Abbasi, N. Islam, and Z. A. Shaikh, 2014: A review of wireless sensors and
25 networks’ applications in agriculture. *Comput. Stand. Interfaces*, doi:10.1016/j.csi.2011.03.004.
- 26 Arndt, C., M. A. Hussain, V. Salvucci, and L. P. Østerdal, 2016: Effects of food price shocks on child
27 malnutrition: The Mozambican experience 2008/2009. *Econ. Hum. Biol.*, **22**, 1–13,
28 doi:10.1016/j.ehb.2016.03.003. <http://www.ncbi.nlm.nih.gov/pubmed/26991234> (Accessed
29 November 3, 2018).
- 30 Arno, A., and S. Thomas, 2016: The efficacy of nudge theory strategies in influencing adult dietary
31 behaviour: a systematic review and meta-analysis. *BMC Public Health*, **16**, 676,
32 doi:10.1186/s12889-016-3272-x.
33 <http://bmcpublichealth.biomedcentral.com/articles/10.1186/s12889-016-3272-x> (Accessed
34 November 5, 2018).
- 35 Arora-Jonsson, S., 2011: Virtue and vulnerability: Discourses on women, gender and climate change.
36 *Glob. Environ. Chang.*, **21**, 744–751, doi:10.1016/J.GLOENVCHA.2011.01.005.
37 <https://www.sciencedirect.com/science/article/pii/S0959378011000069> (Accessed May 3, 2018).
- 38 Aryal, J. P., T. B. Sapkota, M. L. Jat, and D. K. Bishnoi, 2015: On-Farm Economic and
39 Environmental Impact of Zero-Tillage Wheat: a Case of North-West India. *Exp. Agric.*, **51**, 1–
40 16, doi:10.1017/S001447971400012X.
- 41 Aryal, J. P., T. B. Sapkota, C. M. Stirling, M. Jat, H. S. Jat, M. Rai, S. Mittal, and J. M. Sutaliya,
42 2016: Conservation agriculture-based wheat production better copes with extreme climate events
43 than conventional tillage-based systems: A case of untimely excess rainfall in Haryana, India.
44 *Agric. Ecosyst. Environ.*, **233**, 325–335, doi:10.1016/j.agee.2016.09.013.
- 45 Ashoori, D., and M. Sadegh, 2016: Adoption of conservation farming practices for sustainable rice
46 production among small-scale paddy farmers in northern Iran. *Paddy Water Environ.*,
47 doi:10.1007/s10333-016-0543-1.
- 48 Aslam, N. M., and J. Varani, 2016: The Western-Style Diet, Calcium Deficiency and Chronic

- 1 Disease. *J. Nutr. Food Sci.*, **6**, 1–6, doi:10.4172/2155-9600.1000496.
2 [https://www.omicsonline.org/open-access/the-westernstyle-diet-calcium-deficiency-and-](https://www.omicsonline.org/open-access/the-westernstyle-diet-calcium-deficiency-and-chronic-disease-2155-9600-1000496.php?aid=72703)
3 [chronic-disease-2155-9600-1000496.php?aid=72703](https://www.omicsonline.org/open-access/the-westernstyle-diet-calcium-deficiency-and-chronic-disease-2155-9600-1000496.php?aid=72703).
- 4 Asseng, S., and Coauthors, 2015: Rising temperatures reduce global wheat production. *Nat. Clim.*
5 *Chang.*, **5**, 143–147, doi:10.1038/nclimate2470.
6 <http://www.nature.com/doi/finder/10.1038/nclimate2470>.
- 7 Atlin, G. N., J. E. Cairns, and B. Das, 2017: Rapid breeding and varietal replacement are critical to
8 adaptation of cropping systems in the developing world to climate change. *Glob. Food Sec.*, **12**,
9 31–37, doi:10.1016/j.gfs.2017.01.008.
- 10 Audsley, E., M. Brander, J. C. Chatterton, D. Murphy-Bokern, C. Webster, and A. G. Williams, 2010:
11 How low can we go? An assessment of greenhouse gas emissions from the UK foodsystem and
12 the scope reduction by 2050. Report for the WWF and Food ClimateResearch Network.
13 <https://dspace.lib.cranfield.ac.uk/handle/1826/6503> (Accessed April 17, 2019).
- 14 Auffhammer, M., and W. Schlenker, 2014: Empirical studies on agricultural impacts and adaptation.
15 *Energy Econ.*, **46**, 555–561, doi:10.1016/J.ENECO.2014.09.010.
16 <https://www.sciencedirect.com/science/article/pii/S0140988314002205?via%3Dihub> (Accessed
17 March 7, 2019).
- 18 Aune, J. B., A. Coulibaly, and K. E. Giller, 2017: Precision farming for increased land and labour
19 productivity in semi-arid West Africa. A review. *Agron. Sustain. Dev.*, doi:10.1007/s13593-017-
20 0424-z.
- 21 Avetisyan, M., T. Hertel, and G. Sampson, 2014: Is Local Food More Environmentally Friendly? The
22 GHG Emissions Impacts of Consuming Imported versus Domestically Produced Food. *Environ.*
23 *Resour. Econ.*, **58**, 415–462, doi:10.1007/s10640-013-9706-3.
24 <http://link.springer.com/10.1007/s10640-013-9706-3> (Accessed December 6, 2017).
- 25 Awodoyin, R., O. Olubode, J. Ogbu, R. Balogun, J. Nwawuisi, and K. Orji, 2015: Indigenous Fruit
26 Trees of Tropical Africa: Status, Opportunity for Development and Biodiversity Management.
27 *Agric. Sci.*, **6**, 31–41, doi:10.4236/as.2015.61004.
- 28 Ayenew, H. Y., S. Biadgilign, L. Schickramm, G. Abate-Kassa, and J. Sauer, 2018: Production
29 diversification, dietary diversity and consumption seasonality: panel data evidence from Nigeria.
30 *BMC Public Health*, **18**, 988, doi:10.1186/s12889-018-5887-6.
31 <http://www.ncbi.nlm.nih.gov/pubmed/30089472> (Accessed October 22, 2018).
- 32 Ayers, J., and T. Forsyth, 2009: Community-Based Adaptation to Climate Change. *Environ. Sci.*
33 *Policy Sustain. Dev.*, **51**, 22–31, doi:10.3200/ENV.51.4.22-31.
34 <http://www.tandfonline.com/doi/abs/10.3200/ENV.51.4.22-31> (Accessed January 11, 2018).
- 35 Azevedo, A. A., and Coauthors, 2017: Limits of Brazil’s Forest Code as a means to end illegal
36 deforestation. *Proc. Natl. Acad. Sci.*, **114**, 7653–7658, doi:10.1073/pnas.1604768114.
37 <http://www.pnas.org/content/114/29/7653>.
- 38 Backlund, P., A. Janetos, and D. Schimel, 2008: *The effects of climate change on agriculture, land*
39 *resources, water resources, and biodiversity in the United States*.
- 40 Badjeck, M.-C., E. H. Allison, A. S. Halls, and N. K. Dulvy, 2010: Impacts of climate variability and
41 change on fishery-based livelihoods. *Mar. Policy*, **34**, 375–383,
42 doi:10.1016/J.MARPOL.2009.08.007.
43 <https://www.sciencedirect.com/science/article/pii/S0308597X09001237> (Accessed April 12,
44 2019).
- 45 Baglioni, S., B. De Pieri, U. Soler, J. Rosell, and T. Tallarico, 2017: Public Policy Interventions in
46 Surplus Food Recovery in France, Germany, Italy and Spain. *Foodsaving in Europe*, Springer
47 International Publishing, Cham, 37–48 http://link.springer.com/10.1007/978-3-319-56555-2_3
48 (Accessed March 5, 2019).

- 1 Bahrami, H., L. J. De Kok, R. Armstrong, G. J. Fitzgerald, M. Bourgault, S. Henty, M. Tausz, and S.
2 Tausz-posch, 2017: The proportion of nitrate in leaf nitrogen , but not changes in root growth ,
3 are associated with decreased grain protein in wheat under elevated [CO₂]. *J. Plant Physiol.*,
4 **216**, 44–51, doi:10.1016/j.jplph.2017.05.011.
- 5 Bailey, R., and L. Wellesley, 2017: *Chokepoints and Vulnerabilities in Global Food Trade*. Chatham
6 House.,
- 7 ———, and Coauthors, 2015: *Extreme weather and resilience of the global food system. Final Project*
8 *Report from the UK-US Taskforce on Extreme Weather and Global Food System Resilience*.
9 United Kingdom, 17 pp. [http://www.foodsecurity.ac.uk/assets/pdfs/extreme-weather-resilience-](http://www.foodsecurity.ac.uk/assets/pdfs/extreme-weather-resilience-of-global-food-system.pdf)
10 [of-global-food-system.pdf](http://www.foodsecurity.ac.uk/assets/pdfs/extreme-weather-resilience-of-global-food-system.pdf).
- 11 Bajželj, B., K. S. Richards, J. M. Allwood, P. Smith, J. S. Dennis, E. Curmi, and C. A. Gilligan, 2014:
12 Importance of food-demand management for climate mitigation. *Nat. Clim. Chang.*, **4**, 924–929,
13 doi:10.1038/nclimate2353. <http://www.nature.com/doi/10.1038/nclimate2353>.
- 14 Baker, B. H., R. Kröger, J. D. Prevost, T. Pierce, J. J. Ramirez-Avila, J. M. P. Czarnecki, D. Faust,
15 and C. Flora, 2016: A field-scale investigation of nutrient and sediment reduction efficiencies of
16 a low-technology best management practice: Low-grade weirs. *Ecol. Eng.*, **91**, 240–248,
17 doi:10.1016/j.ecoleng.2016.02.038. <http://dx.doi.org/10.1016/j.ecoleng.2016.02.038>.
- 18 Baldermann, S., and Coauthors, 2016: Are Neglected Plants the Food for the Future? *CRC. Crit. Rev.*
19 *Plant Sci.*, **35**, 106–119, doi:10.1080/07352689.2016.1201399.
20 <https://www.tandfonline.com/doi/full/10.1080/07352689.2016.1201399> (Accessed October 21,
21 2018).
- 22 Baldos, U. L. C., and T. W. Hertel, 2015: The role of international trade in managing food security
23 risks from climate change. *Food Secur.*, **7**, 275–290, doi:10.1007/s12571-015-0435-z.
- 24 Balmford, A., and Coauthors, 2018: The environmental costs and benefits of high-yield farming. *Nat.*
25 *Sustain.*, doi:10.1038/s41893-018-0138-5.
- 26 Balzan, M. V., G. Bocci, and A. C. Moonen, 2016: Utilisation of plant functional diversity in
27 wildflower strips for the delivery of multiple agroecosystem services. *Entomol. Exp. Appl.*, **158**,
28 304–319, doi:10.1111/eea.12403.
- 29 BAPEN, 2012: *Nutrition screening survey in the uk and republic of ireland in 2011*. Redditch, UK, 1-
30 76 pp.
- 31 Barange, M., T. Bahri, M. C. M. Beveridge, K. L. Cochrane, S. Funge-Smith, and F. Poulain, 2018:
32 *Impacts of climate change on fisheries and aquaculture: synthesis of current knowledge,*
33 *adaptation and mitigation options. FAO Fisheries and Aquaculture Technical Paper No. 627.*
34 Rome, Italy, 628 pp.
- 35 Barrett, C. B., 2006: *Food Aid's Intended and Unintended Consequences*. Rome, Italy, 27 pp.
36 <http://www.fao.org/3/a-ag301t.pdf>.
- 37 Barriopedro, D., E. M. Fischer, J. Luterbacher, R. M. Trigo, and R. García-Herrera, 2011: The Hot
38 Summer of 2010: Redrawing the Temperature Record Map of Europe. *Science (80-.)*, **332**,
39 220–224, doi:10.1126/science.1201224.
- 40 Bartlett, S., 2008: Climate change and urban children: impacts and implications for adaptation in low-
41 and middle-income countries. *Environ. Urban.*, **20**, 501–519, doi:10.1177/0956247808096125.
42 <http://journals.sagepub.com/doi/10.1177/0956247808096125> (Accessed May 25, 2018).
- 43 Batalla, I., M. T. Knudsen, L. Mogensen, Ó. del Hierro, M. Pinto, and J. E. Hermansen, 2015: Carbon
44 footprint of milk from sheep farming systems in Northern Spain including soil carbon
45 sequestration in grasslands. *J. Clean. Prod.*, **104**, 121–129,
46 doi:10.1016/J.JCLEPRO.2015.05.043.
47 <https://www.sciencedirect.com/science/article/pii/S0959652615005909> (Accessed April 17,
48 2019).

- 1 Bates, B., C. Roberts, H. Lepps, and L. Porter, 2017: *The Food and You Survey Wave 4 Combined*
2 *Report for England, Wales and Northern Ireland.*
3 [https://www.food.gov.uk/sites/default/files/media/document/food-and-you-w4-combined-](https://www.food.gov.uk/sites/default/files/media/document/food-and-you-w4-combined-report_0.pdf)
4 [report_0.pdf](https://www.food.gov.uk/sites/default/files/media/document/food-and-you-w4-combined-report_0.pdf) (Accessed October 8, 2018).
- 5 Batima, P., L. Natsagdorj, and N. Batnasan, 2008: Vulnerability of Mongolia's Pastoralists to Climate
6 Extremes and Changes. *Climate Change and Vulnerability and Adaptation*, N. Leary, C. Conde,
7 J. Kulkarni, A. Nyong, and J. Pulin, Eds., Earthscan/Taylor and Francis, London, 67–87
8 [https://books.google.es/books?hl=es&lr=&id=yUb5AQAAQBAJ&oi=fnd&pg=PA67&dq=pasto-](https://books.google.es/books?hl=es&lr=&id=yUb5AQAAQBAJ&oi=fnd&pg=PA67&dq=pastoralist+climate+change&ots=tuXQQ4agcx&sig=c45GUR0Ak0OtBBGFu4cWfStCxck#v=onepage&q=pastoralist+climate+change&f=false)
9 [ralist+climate+change&ots=tuXQQ4agcx&sig=c45GUR0Ak0OtBBGFu4cWfStCxck#v=onepa-](https://books.google.es/books?hl=es&lr=&id=yUb5AQAAQBAJ&oi=fnd&pg=PA67&dq=pastoralist+climate+change&ots=tuXQQ4agcx&sig=c45GUR0Ak0OtBBGFu4cWfStCxck#v=onepage&q=pastoralist+climate+change&f=false)
10 [ge&q=pastoralist climate change&f=false](https://books.google.es/books?hl=es&lr=&id=yUb5AQAAQBAJ&oi=fnd&pg=PA67&dq=pastoralist+climate+change&ots=tuXQQ4agcx&sig=c45GUR0Ak0OtBBGFu4cWfStCxck#v=onepage&q=pastoralist+climate+change&f=false) (Accessed March 15, 2019).
- 11 Battersby, J., 2012: Beyond the food desert: finding ways to speak about urban food security in south
12 africa. *Geogr. Ann. Ser. B, Hum. Geogr.*, **94**, 141–159, doi:10.1111/j.1468-0467.2012.00401.x.
13 <https://www.tandfonline.com/doi/full/10.1111/j.1468-0467.2012.00401.x> (Accessed March 20,
14 2019).
- 15 Battilani, P., and Coauthors, 2016: Aflatoxin B1 contamination in maize in Europe increases due to
16 climate change. *Sci. Rep.*, **6**, 24328, doi:10.1038/srep24328.
17 <http://www.nature.com/articles/srep24328> (Accessed January 8, 2018).
- 18 Baynes, J., J. Herbohn, and W. Unsworth, 2017: Reforesting the grasslands of Papua New Guinea:
19 The importance of a family-based approach. *J. Rural Stud.*, **56**, 124–131,
20 doi:10.1016/J.JRURSTUD.2017.09.012.
- 21 Bebbler, D. P., 2015: Range-Expanding Pests and Pathogens in a Warming World. *Annual Review of*
22 *Phytopathology*, Vol 53, N.K. VanAlfen, Ed., Vol. 53 of, 335–356.
- 23 ———, and S. J. Gurr, 2015: Crop-destroying fungal and oomycete pathogens challenge food security.
24 *Fungal Genet. Biol.*, **74**, 62–64, doi:10.1016/j.fgb.2014.10.012.
- 25 ———, T. Holmes, and S. J. Gurr, 2014: The global spread of crop pests and pathogens. *Glob. Ecol.*
26 *Biogeogr.*, **23**, 1398–1407, doi:10.1111/geb.12214.
- 27 Bee, B. A., 2016: Power, perception, and adaptation: Exploring gender and social–environmental risk
28 perception in northern Guanajuato, Mexico. *Geoforum*, **69**, 71–80,
29 doi:10.1016/J.GEOFORUM.2015.12.006.
30 <https://www.sciencedirect.com/science/article/pii/S0016718515301548?via%3Dihub> (Accessed
31 October 29, 2018).
- 32 Beheshti, R., J. C. Jones-Smith, and T. Igusa, 2017: Taking dietary habits into account: A
33 Computational method for modeling food choices that goes beyond price. *PLoS One*, **12**, 1–13,
34 doi:10.1371/journal.pone.0178348.
- 35 Beletse, Y. G., and Coauthors, 2015: Projected Impacts of Climate Change Scenarios on the
36 Production of Maize in Southern Africa: An Integrated Assessment Case Study of the
37 Bethlehem District, Central Free State, South Africa. *Handbook of Climate Change and*
38 *agroecosystems: The Agricultural Model Intercomparison and Improvement Project Integrated*
39 *Crop and Economic Assessments, Part 2*, World Scientific, 125–157.
- 40 Ben-Ari, T., J. Boé, P. Ciais, R. Lecerf, M. Van der Velde, and D. Makowski, 2018: Causes and
41 implications of the unforeseen 2016 extreme yield loss in the breadbasket of France. *Nat.*
42 *Commun.*, **9**, 1627, doi:10.1038/s41467-018-04087-x.
43 <http://www.ncbi.nlm.nih.gov/pubmed/29691405> (Accessed April 4, 2019).
- 44 Benjaminsen, T. A., K. Alinon, H. Buhaug, and J. T. Buseeth, 2012: Does climate change drive land-
45 use conflicts in the Sahel? *J. Peace Res.*, **49**, 97–111, doi:10.1177/0022343311427343.
46 <http://journals.sagepub.com/doi/10.1177/0022343311427343> (Accessed March 15, 2019).
- 47 Benkeblia, N., 2018: *Climate Change and Crop Production. Foundations for Agroecosystem*
48 *Resilience*. N. Benkeblia, Ed. CRC Press, Boca Raton, FL : Taylor & Francis, 2018. | Series:
49 *Advances in agroecology*, 207 pp. <https://www.taylorfrancis.com/books/9781315391854>

- 1 (Accessed March 9, 2019).
- 2 Benton, T. G., and T. Bailey, 2019: The paradox of productivity: agricultural productivity promotes
3 food system inefficiency. *Glob. Sustain.*, **In Press**.
- 4 Benton, T. G., R. Bailey, A. Froggatt, R. King, B. Lee, and L. Wellesley, 2018: Designing sustainable
5 landuse in a 1.5° world: the complexities of projecting multiple ecosystem services from land.
6 *Curr. Opin. Environ. Sustain.*, **31**, 88–95, doi:<https://doi.org/10.1016/j.cosust.2018.01.011>.
- 7 Beringer, T., W. Lucht, and S. Schaphoff, 2011: Bioenergy production potential of global biomass
8 plantations under environmental and agricultural constraints. *GCB Bioenergy*, **3**, 299–312,
9 doi:10.1111/j.1757-1707.2010.01088.x. <http://doi.wiley.com/10.1111/j.1757-1707.2010.01088.x>
10 (Accessed April 17, 2019).
- 11 Berke, P. R., and M. R. Stevens, 2016: Land Use Planning for Climate Adaptation : Theory and
12 Practice. *J. Plan. Educ. Res.*, **36**, doi:10.1177/0739456X16660714.
- 13 Bernstein, L. E., E. T. Auer, M. Wagner, and C. W. Ponton, 2008: Spatiotemporal dynamics of
14 audiovisual speech processing. *Neuroimage*, **39**, 423–435,
15 doi:10.1016/j.neuroimage.2007.08.035.
- 16 Bett, B., P. Kiunga, J. Gachohi, C. Sindato, D. Mbotha, T. Robinson, J. Lindahl, and D. Grace, 2017:
17 Effects of climate change on the occurrence and distribution of livestock diseases. *Prev. Vet.
18 Med.*, **137**, 119–129, doi:10.1016/j.prevetmed.2016.11.019.
19 <https://www.sciencedirect.com/science/article/pii/S0167587716306316> (Accessed October 20,
20 2018).
- 21 Betts, R. A., and Coauthors, 2018: Changes in climate extremes, fresh water availability and
22 vulnerability to food insecurity projected at 1.5°C and 2°C global warming with a higher-
23 resolution global climate model. *Philos. Trans. R. Soc. A Math. Eng. Sci.*, **376**, 20160452,
24 doi:10.1098/rsta.2016.0452.
25 <http://rsta.royalsocietypublishing.org/lookup/doi/10.1098/rsta.2016.0452> (Accessed April 16,
26 2019).
- 27 Beuchle, R., R. C. Grecchi, Y. E. Shimabukuro, R. Seliger, H. D. Eva, E. Sano, and F. Achard, 2015:
28 Land cover changes in the Brazilian Cerrado and Caatinga biomes from 1990 to 2010 based on a
29 systematic remote sensing sampling approach. *Appl. Geogr.*, **58**, 116–127,
30 doi:10.1016/J.APGEOG.2015.01.017.
31 <https://www.sciencedirect.com/science/article/pii/S0143622815000284?via%3Dihub> (Accessed
32 May 18, 2018).
- 33 Bhat, Z. F., S. Kumar, and H. Fayaz, 2015: In vitro meat production: Challenges and benefits over
34 conventional meat production. *J. Integr. Agric.*, **14**, 241–248, doi:10.1016/S2095-
35 3119(14)60887-X. <https://www.sciencedirect.com/science/article/pii/S209531191460887X>
36 (Accessed November 4, 2018).
- 37 Bhatta, G. D., and P. K. Aggarwal, 2016: Coping with weather adversity and adaptation to climatic
38 variability: a cross-country study of smallholder farmers in South Asia. *Clim. Dev.*, **8**, 145–157,
39 doi:10.1080/17565529.2015.1016883. <https://doi.org/10.1080/17565529.2015.1016883>.
- 40 Bhattacharyya, P. N., M. P. Goswami, and L. H. Bhattacharyya, 2016: Perspective of benefi cial
41 microbes in agriculture under changing climatic scenario : A review. *J. Phytol.*, **8**, 26–41,
42 doi:10.19071/jp.2016.v8.3022.
- 43 Biesbroek, S., H. B. Bueno-de-Mesquita, P. H. M. Peeters, W. M. Verschuren, Y. T. van der Schouw,
44 G. F. H. Kramer, M. Tyszler, and E. H. M. Temme, 2014: Reducing our environmental footprint
45 and improving our health: greenhouse gas emission and land use of usual diet and mortality in
46 EPIC-NL: a prospective cohort study. *Environ. Health*, **13**, 27, doi:10.1186/1476-069X-13-27.
47 <http://www.scopus.com/inward/record.url?eid=2-s2.0-84907670666&partnerID=tZOtx3y1>.
- 48 Biggs, E. M., E. L. Tompkins, J. Allen, C. Moon, and R. Allen, 2013: Agricultural adaptation to
49 climate change: observations from the Mid-Hills of Nepal. *Clim. Dev.*, **5**, 165–173,

- 1 doi:10.1080/17565529.2013.789791.
2 <http://www.tandfonline.com/doi/abs/10.1080/17565529.2013.789791> (Accessed November 1,
3 2018).
- 4 Bioversity International, 2016: *Mainstreaming Agrobiodiversity in Sustainable Food Systems:
5 Scientific Foundations for an Agrobiodiversity Index-Summary*. 30 pp.
- 6 Birney, C. I., K. F. Franklin, F. T. Davidson, and M. E. Webber, 2017: An assessment of individual
7 footprints attributed to diets and food waste in the United States. *Environ. Res. Lett.*, **12**,
8 105008, doi:10.1088/1748-9326/aa8494. <http://stacks.iop.org/1748-9326/12/i=10/a=105008>.
- 9 Bisbis, M. B., N. Gruda, and M. Blanke, 2018: Potential impacts of climate change on vegetable
10 production and product quality - A review. *J. Clean. Prod.*, **170**, 1602–1620,
11 doi:10.1016/j.jclepro.2017.09.224.
- 12 Blanchard, J. L., and Coauthors, 2017: Linked sustainability challenges and trade-offs among
13 fisheries, aquaculture and agriculture. *Nat. Ecol. Evol.*, **1**, 1240–1249, doi:10.1038/s41559-017-
14 0258-8.
- 15 Blanco Sepúlveda, R., and A. Aguilar Carrillo, 2016: The erosion threshold for a sustainable
16 agriculture in cultures of bean (*Phaseolus vulgaris* L.) under conventional tillage and no-tillage
17 in Northern Nicaragua. *Soil Use Manag.*, **32**, 368–380, doi:10.1111/sum.12271.
18 <http://doi.wiley.com/10.1111/sum.12271> (Accessed April 24, 2019).
- 19 Bloom, A. J., M. Burger, B. A. Kimball, and P. J. Pinter, 2014: Nitrate assimilation is inhibited by
20 elevated CO₂ in field-grown wheat. *Nat. Clim. Chang.*, **4**, 477–480, doi:10.1038/nclimate2183.
- 21 Bloom, D. E., S. Chatterji, P. Kowal, P. Lloyd-Sherlock, M. McKee, B. Rechel, L. Rosenberg, and J.
22 P. Smith, 2015: Macroeconomic implications of population ageing and selected policy
23 responses. *Lancet*, **385**, 649–657.
- 24 Bobojonov, I., and A. Aw-Hassan, 2014: Impacts of climate change on farm income security in
25 Central Asia: An integrated modeling approach. *Agric. Ecosyst. Environ.*, **188**, 245–255,
26 doi:10.1016/j.agee.2014.02.033. <http://dx.doi.org/10.1016/j.agee.2014.02.033>.
- 27 Bocco, G., and B. M. Napoletano, 2017: The prospects of terrace agriculture as an adaptation to
28 climate change in Latin America. *Geogr. Compass*, **11**, 1–13, doi:10.1111/gec3.12330.
- 29 Bodin, P., S. Olin, T. a. M. Pugh, and a. Arneeth, 2016: Accounting for interannual variability in
30 agricultural intensification: The potential of crop selection in Sub-Saharan Africa. *Agric. Syst.*,
31 **148**, 159–168, doi:10.1016/j.agry.2016.07.012.
- 32 de Boer, I., and Coauthors, 2011: Greenhouse gas mitigation in animal production: towards an
33 integrated life cycle sustainability assessment. *Curr. Opin. Environ. Sustain.*, **3**, 423–431,
34 doi:10.1016/J.COSUST.2011.08.007.
- 35 de Boer, J., H. Schösler, and H. Aiking, 2014: “Meatless days” or “less but better”? Exploring
36 strategies to adapt Western meat consumption to health and sustainability challenges. *Appetite*,
37 **76**, 120–128, doi:10.1016/J.APPET.2014.02.002.
38 <https://www.sciencedirect.com/science/article/pii/S0195666314000907> (Accessed April 8,
39 2019).
- 40 Boerema, A., A. Peeters, S. Swolfs, F. Vandevenne, S. Jacobs, J. Staes, and P. Meire, 2016: Soybean
41 Trade: Balancing Environmental and Socio-Economic Impacts of an Intercontinental Market.
42 *PLoS One*, **11**, e0155222, doi:10.1371/journal.pone.0155222.
43 <https://dx.plos.org/10.1371/journal.pone.0155222> (Accessed November 5, 2018).
- 44 Boettcher, P. J., and Coauthors, 2015: Genetic resources and genomics for adaptation of livestock to
45 climate change STRATEGIES. *Front. Genet.*, **5**, 2014–2016, doi:10.3389/fgene.2014.00461.
- 46 Boetto, H., and J. McKinnon, 2013: Rural Women and Climate Change: A Gender-inclusive
47 Perspective. *Aust. Soc. Work*, **66**, 234–247, doi:10.1080/0312407X.2013.780630.
48 <http://www.tandfonline.com/doi/abs/10.1080/0312407X.2013.780630> (Accessed March 20,

- 1 2019).
- 2 Bogaerts, M., L. Cirhigiri, I. Robinson, M. Rodkin, R. Hajjar, C. Costa Junior, and P. Newton, 2017:
3 Climate change mitigation through intensified pasture management: Estimating greenhouse gas
4 emissions on cattle farms in the Brazilian Amazon. *J. Clean. Prod.*, **162**, 1539–1550,
5 doi:10.1016/J.JCLEPRO.2017.06.130.
6 <https://www.sciencedirect.com/science/article/pii/S0959652617313008> (Accessed May 18,
7 2018).
- 8 van den Bold, M., A. Dillon, D. Olney, M. Ouedraogo, A. Pedehombga, and A. Quisumbing, 2015:
9 Can Integrated Agriculture-Nutrition Programmes Change Gender Norms on Land and Asset
10 Ownership? Evidence from Burkina Faso. *J. Dev. Stud.*, **51**, 1155–1174,
11 doi:10.1080/00220388.2015.1036036.
12 <http://www.tandfonline.com/doi/full/10.1080/00220388.2015.1036036> (Accessed October 29,
13 2018).
- 14 Bond, H., and Coauthors, 2018: *Developing a Global Compendium on Water Quality Guidelines*. 1-
15 79 pp. <https://www.iwra.org/wp-content/uploads/2018/11/WQ-compendium-final-1.pdf>
16 (Accessed April 24, 2019).
- 17 Boone, R. B., R. T. Conant, J. Sircely, P. K. Thornton, and M. Herrero, 2017: Climate Change
18 Impacts on Selected Global Rangeland Ecosystem Services. *Glob. Chang. Biol.*,
19 doi:10.1111/gcb.13995. <http://doi.wiley.com/10.1111/gcb.13995> (Accessed December 21,
20 2017).
- 21 Boucher, D., and D. Chi, 2018: Amazon Deforestation in Brazil: What Has Not Happened and How
22 the Global Media Covered It. *Trop. Conserv. Sci.*, **11**, 1940082918794325,
23 doi:10.1177/1940082918794325. <https://doi.org/10.1177/1940082918794325>.
- 24 Bouroncle, C., P. Imbach, B. Rodríguez-Sánchez, C. Medellín, A. Martínez-Valle, and P. Läderach,
25 2017: Mapping climate change adaptive capacity and vulnerability of smallholder agricultural
26 livelihoods in Central America: ranking and descriptive approaches to support adaptation
27 strategies. *Clim. Change*, **141**, 123–137, doi:10.1007/s10584-016-1792-0.
28 <http://link.springer.com/10.1007/s10584-016-1792-0> (Accessed April 16, 2019).
- 29 Bowles, N., S. Alexander, and M. Hadjikakou, 2019: The livestock sector and planetary boundaries:
30 A “limits to growth” perspective with dietary implications. *Ecol. Econ.*, **160**, 128–136,
31 doi:10.1016/J.ECOLECON.2019.01.033.
- 32 Brandt, M., and Coauthors, 2018: Reduction of tree cover in West African woodlands and promotion
33 in semi-arid farmlands. *Nat. Geosci.*, **11**, 328–333, doi:10.1038/s41561-018-0092-x.
34 <https://doi.org/10.1038/s41561-018-0092-x>.
- 35 Brashares, J. S., P. Arcese, M. K. Sam, P. B. Coppolillo, a R. E. Sinclair, and A. Balmford, 2004:
36 Bushmeat hunting, wildlife declines, and fish supply in West Africa. *Science (80-.)*, **306**, 1180–
37 1183, doi:10.1126/science.1102425.
- 38 Breisinger, C., O. Ecker, J.-F. Trinh Tan, and others, 2015: Conflict and food insecurity: How do we
39 break the links? *IFPRI B. chapters*, 51–60.
- 40 Bren d’Amour, C., and Coauthors, 2017: Future urban land expansion and implications for global
41 croplands. *Proc. Natl. Acad. Sci.*, **114**, 8939 LP-8944.
42 <http://www.pnas.org/content/114/34/8939.abstract>.
- 43 Brodt, S., K. J. Kramer, A. Kendall, and G. Feenstra, 2013: Comparing environmental impacts of
44 regional and national-scale food supply chains: A case study of processed tomatoes. *Food*
45 *Policy*, **42**, 106–114, doi:https://doi.org/10.1016/j.foodpol.2013.07.004.
46 <http://www.sciencedirect.com/science/article/pii/S0306919213001012>.
- 47 Brooks, J., and A. Matthews, 2015: Trade Dimensions of Food Security. *OECD Food, Agric. Fish.*
48 *Pap.*, 1–45, doi:10.1787/5js65xn790nv-en.

- 1 Brown, B., I. Nuberg, and R. Llewellyn, 2017a: Negative evaluation of conservation agriculture:
2 perspectives from African smallholder farmers. *Int. J. Agric. Sustain.*, **15**, 467–481,
3 doi:10.1080/14735903.2017.1336051.
4 <https://www.tandfonline.com/doi/full/10.1080/14735903.2017.1336051> (Accessed October 29,
5 2018).
- 6 Brown, M. E., 2016: Remote sensing technology and land use analysis in food security assessment
7 security assessment. **4248**, doi:10.1080/1747423X.2016.1195455.
- 8 Brown, M. E., E. R. Carr, K. L. Grace, K. Wiebe, C. C. Funk, W. Attavanich, P. Backlund, and L.
9 Buja, 2017b: Do markets and trade help or hurt the global food system adapt to climate change?
10 *Food Policy*, **68**, 154–159, doi:10.1016/j.foodpol.2017.02.004.
11 <http://dx.doi.org/10.1016/j.foodpol.2017.02.004>.
- 12 Brunner, S. H., and A. Grêt-Regamey, 2016: Policy strategies to foster the resilience of mountain
13 social-ecological systems under uncertain global change. *Environ. Sci. Policy*, **66**, 129–139,
14 doi:10.1016/j.envsci.2016.09.003. <http://dx.doi.org/10.1016/j.envsci.2016.09.003>.
- 15 Bryan, E., and J. A. Behrman, 2013: *Community-based adaptation to climate change: A theoretical*
16 *framework, overview of key issues and discussion of gender differentiated priorities and*
17 *participation*. Washington, D.C.,.
- 18 Bryngelsson, D., F. Hedenus, D. Johansson, C. Azar, and S. Wirsenius, 2017: How Do Dietary
19 Choices Influence the Energy-System Cost of Stabilizing the Climate? *Energies*, **10**, 182,
20 doi:10.3390/en10020182. <http://www.mdpi.com/1996-1073/10/2/182>.
- 21 Buckwell, A., and Coauthors, 2014: *The Sustainable Intensification of European Agriculture: A*
22 *review sponsored by The Rise Foundation*. 98 pp.
- 23 Buggy, L., and K. E. McNamara, 2016: The need to reinterpret “community” for climate change
24 adaptation: a case study of Pele Island, Vanuatu. *Clim. Dev.*, **8**, 270–280,
25 doi:10.1080/17565529.2015.1041445.
- 26 Buhaug, H., 2016: Climate change and conflict: taking stock. *Peace Econ. Peace Sci. Public Policy*,
27 **22**, 331–338.
- 28 Bungenstab, D. J., 2012: *Pecuária de corte brasileira: redução do aquecimento global pela eficiência*
29 *dos sistemas de produção*. Brasília, Brasil: Embrapa Gado de Corte.,.
- 30 Burney, J., and V. Ramanathan, 2014: Recent climate and air pollution impacts on Indian agriculture.
31 doi:10.1073/pnas.1317275111. <http://www.pnas.org/content/pnas/111/46/16319.full.pdf>
32 (Accessed May 6, 2018).
- 33 Burnham, M., and Z. Ma, 2016: Linking smallholder farmer climate change adaptation decisions to
34 development. *Clim. Dev.*, **8**, 289–311, doi:10.1080/17565529.2015.1067180.
35 <https://doi.org/10.1080/17565529.2015.1067180>.
- 36 Burns, W., and S. Nicholson, 2017: Bioenergy and carbon capture with storage (BECCS): the
37 prospects and challenges of an emerging climate policy response. *J. Environ. Stud. Sci.*, **7**, 527–
38 534, doi:10.1007/s13412-017-0445-6. <http://link.springer.com/10.1007/s13412-017-0445-6>.
- 39 Byerlee, D., J. Stevenson, and N. Villoria, 2014: Does intensification slow crop land expansion or
40 encourage deforestation? *Glob. Food Sec.*, **3**, 92–98, doi:10.1016/J.GFS.2014.04.001.
41 <https://www.sciencedirect.com/science/article/pii/S221191241400011X> (Accessed March 8,
42 2018).
- 43 CABI, 2019: About Plantwise. <https://www.plantwise.org/about-plantwise/> (Accessed March 5,
44 2019).
- 45 Caldeira, S., S. S. Bonsmann, I. Bakogianni, C. Gauci, A. Calleja, and A. Furtado, 2017: *Public*
46 *Procurement of Food for Health*. 86 pp. [https://ec.europa.eu/jrc/sites/jrcsh/files/public-](https://ec.europa.eu/jrc/sites/jrcsh/files/public-procurement-food-health-technical-report.pdf)
47 [procurement-food-health-technical-report.pdf](https://ec.europa.eu/jrc/sites/jrcsh/files/public-procurement-food-health-technical-report.pdf).

- 1 Calori, A., and A. Magarini, 2015: *Food and the cities: politiche del cibo per città sostenibili*.
2 Edizioni Ambiente,.
- 3 Calvet-Mir, L., and Coauthors, 2018: The Contribution of Traditional Agroecological Knowledge as a
4 Digital Commons to Agroecological Transitions: The Case of the CONECT-e Platform.
5 *Sustainability*, **10**, 3214, doi:10.3390/su10093214. <http://www.mdpi.com/2071-1050/10/9/3214>
6 (Accessed March 2, 2019).
- 7 Caminade, C., J. van Dijk, M. Baylis, and D. Williams, 2015: Modelling recent and future climatic
8 suitability for fasciolosis in Europe. *Geospat. Health*, **2**, 301–308.
- 9 Cammarano, D., and Coauthors, 2016: Uncertainty of wheat water use: Simulated patterns and
10 sensitivity to temperature and CO₂. *F. Crop. Res.*, **198**, 80–92, doi:10.1016/j.fcr.2016.08.015.
11 <http://dx.doi.org/10.1016/j.fcr.2016.08.015>.
- 12 Campbell, B. M., and Coauthors, 2016: Reducing risks to food security from climate change. *Glob.*
13 *Food Sec.*, **11**, 34–43, doi:10.1016/j.gfs.2016.06.002.
14 <http://linkinghub.elsevier.com/retrieve/pii/S2211912415300262>.
- 15 Campbell, J., R. Bedford, and R. Bedford, 2014: Migration and Climate Change in Oceania. *People*
16 *on the Move in a Changing Climate: The Regional Impact of Environmental Change on*
17 *Migration*, E. Piguet and F. Laczko, Eds., Springer Netherlands, Dordrecht, 177–204
18 https://doi.org/10.1007/978-94-007-6985-4_8.
- 19 Campbell, J. R., 2014: Climate-Change Migration in the Pacific. *Contemp. Pac.*, **26**, 1–28,
20 doi:10.1353/cp.2014.0023.
21 http://muse.jhu.edu/content/crossref/journals/contemporary_pacific/v026/26.1.campbell.html.
- 22 Campbell, J. R., 2015: Development, global change and traditional food security in Pacific Island
23 countries. *Reg. Environ. Chang.*, **15**, 1313–1324, doi:10.1007/s10113-014-0697-6.
- 24 Campillo, G., M. Mullan, and L. Vallejo, 2017: Climate Change Adaptation and Financial Protection.
25 *OECD Environ. Work. Pap.*,
- 26 Canto, T., M. A. Aranda, and A. Fereres, 2009: Climate change effects on physiology and population
27 processes of hosts and vectors that influence the spread of hemipteran-borne plant viruses. *Glob.*
28 *Chang. Biol.*, **15**, 1884–1894, doi:10.1111/j.1365-2486.2008.01820.x.
- 29 Cardoso, A. S., A. Berndt, A. Leytem, B. J. R. Alves, I. das N. O. de Carvalho, L. H. de Barros
30 Soares, S. Urquiaga, and R. M. Boddey, 2016: Impact of the intensification of beef production in
31 Brazil on greenhouse gas emissions and land use. *Agric. Syst.*, **143**, 86–96,
32 doi:10.1016/j.agsy.2015.12.007.
33 <https://linkinghub.elsevier.com/retrieve/pii/S0308521X15300652>.
- 34 Carlson, K. M., and Coauthors, 2017: Greenhouse gas emissions intensity of global croplands. *Nat.*
35 *Clim. Chang.*, **7**, 63–72, doi:10.1038/nclimate3158.
- 36 Carvalho, P. C. de F., and B. Dedieu, 2014: Integrated crop–livestock systems: Strategies to achieve
37 synergy between agricultural production and environmental quality. *Agric. Ecosyst. Environ.*,
38 **190**, 4–8, doi:10.1016/J.AGEE.2013.08.009.
- 39 Carvalho, R., M. Adami, S. Amaral, F. G. Bezerra, and A. P. D. de Aguiar, 2019: Changes in
40 secondary vegetation dynamics in a context of decreasing deforestation rates in Pará, Brazilian
41 Amazon. *Appl. Geogr.*, **106**, 40–49, doi:10.1016/J.APGEOG.2019.03.001.
- 42 Casado-Vara, R., J. Prieto, F. De la Prieta, and J. M. Corchado, 2018: How blockchain improves the
43 supply chain: case study alimentary supply chain. *Procedia Comput. Sci.*, **134**, 393–398,
44 doi:10.1016/J.PROCS.2018.07.193.
45 <https://www.sciencedirect.com/science/article/pii/S187705091831158X> (Accessed April 17,
46 2019).
- 47 Castañé, S., and A. Antón, 2017: Assessment of the nutritional quality and environmental impact of
48 two food diets: A Mediterranean and a vegan diet. *J. Clean. Prod.*,

- 1 doi:10.1016/j.jclepro.2017.04.121.
2 <http://linkinghub.elsevier.com/retrieve/pii/S095965261730848X>.
- 3 Cawthorn, D.-M., and L. C. Hoffman, 2015: The bushmeat and food security nexus: A global account
4 of the contributions, conundrums and ethical collisions. *Food Res. Int.*, **v. 76**, 906-925–2015
5 v.76, doi:10.1016/j.foodres.2015.03.025.
- 6 CCAFS, 2012: Agriculture’s role in both adaptation and mitigation discussed during climate talks.
- 7 CDP, 2018: Tracking climate progress 2017 - CDP. [https://www.cdp.net/en/research/global-](https://www.cdp.net/en/research/global-reports/tracking-climate-progress-2017)
8 [reports/tracking-climate-progress-2017](https://www.cdp.net/en/research/global-reports/tracking-climate-progress-2017) (Accessed May 30, 2018).
- 9 —, 2019: Carbon Disclosure Project. <https://www.cdp.net/en> (Accessed April 24, 2019).
- 10 Ceddia, M. G., S. Sedlacek, N. O. Bardsley, and S. Gomez-y-Paloma, 2013: Sustainable agricultural
11 intensification or Jevons paradox? The role of public governance in tropical South America.
12 *Glob. Environ. Chang.*, **23**, 1052–1063, doi:10.1016/j.gloenvcha.2013.07.005.
13 <https://linkinghub.elsevier.com/retrieve/pii/S0959378013001118>.
- 14 Ceglar, A., A. Toreti, C. Prodhomme, M. Zampieri, M. Turco, and F. J. Doblas-Reyes, 2018: Land-
15 surface initialisation improves seasonal climate prediction skill for maize yield forecast. *Sci.*
16 *Rep.*, **8**, 1322, doi:10.1038/s41598-018-19586-6. <https://doi.org/10.1038/s41598-018-19586-6>.
- 17 Celentano, D., and Coauthors, 2017: Towards zero deforestation and forest restoration in the Amazon
18 region of Maranhão state, Brazil. *Land use policy*, **68**, 692–698,
19 doi:10.1016/J.LANDUSEPOL.2017.07.041.
20 <https://www.sciencedirect.com/science/article/pii/S0264837717302296?via%3Dihub> (Accessed
21 May 18, 2018).
- 22 Chabejong, N. E., 2016: A Review on the Impact of Climate Change on Food Security and
23 Malnutrition in the Sahel Region of Cameroon. *Climate Change and Health*, Springer, 133–148.
- 24 Chadwick, V., 2017: Banks to incentivize sustainable supply chains under DFID-backed pilot |
25 Devex. [https://www.devex.com/news/banks-to-incentivize-sustainable-supply-chains-under-](https://www.devex.com/news/banks-to-incentivize-sustainable-supply-chains-under-dfid-backed-pilot-91730)
26 [dfid-backed-pilot-91730](https://www.devex.com/news/banks-to-incentivize-sustainable-supply-chains-under-dfid-backed-pilot-91730) (Accessed April 24, 2019).
- 27 Chalak, A., C. Abou-Daher, J. Chaaban, and M. G. Abiad, 2016: The global economic and regulatory
28 determinants of household food waste generation: A cross-country analysis. *Waste Manag.*, **48**,
29 418–422, doi:10.1016/J.WASMAN.2015.11.040.
30 <https://www.sciencedirect.com/science/article/pii/S0956053X15302221> (Accessed November 5,
31 2018).
- 32 Chalise, S., and A. Naranpanawa, 2016: Climate change adaptation in agriculture: A computable
33 general equilibrium analysis of land-use change in Nepal. *Land use policy*, **59**, 241–250,
34 doi:10.1016/j.landusepol.2016.09.007. <http://dx.doi.org/10.1016/j.landusepol.2016.09.007>.
- 35 Challinor, A., W. N. Adger, T. G. Benton, D. Conway, M. Joshi, and D. Frame, 2018: Transmission
36 of climate risks across sectors and borders. *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*,
- 37 Challinor, A. J., and Coauthors, 2016: *UK Climate Change Risk Assessment Evidence Report:*
38 *Chapter 7, International Dimensions. Report prepared for the Adaptation Sub-Committee of the*
39 *Committee on Climate Change*. London, UK, 85 pp pp.
- 40 Charan, D., M. Kaur, and P. Singh, 2017: Customary Land and Climate Change Induced Relocation—
41 A Case Study of Vunidogoloa Village, Vanua Levu, Fiji. *Climate Change Adaptation in Pacific*
42 *Countries*, 19–34 <http://link.springer.com/10.1007/978-3-319-50094-2>.
- 43 Charles, A., P. Rochette, J. K. Whalen, D. A. Angers, M. H. Chantigny, and N. Bertrand, 2017:
44 Global nitrous oxide emission factors from agricultural soils after addition of organic
45 amendments: A meta-analysis. *Agric. Ecosyst. Environ.*, **236**, 88–98,
46 doi:10.1016/J.AGEE.2016.11.021.
47 <https://www.sciencedirect.com/science/article/pii/S0167880916305795> (Accessed May 25,
48 2018).

- 1 Chartzoulakis, K., and M. Bertaki, 2015: Sustainable Water Management in Agriculture under
2 Climate Change. *Agric. Agric. Sci. Procedia*, **4**, 88–98, doi:10.1016/j.aaspro.2015.03.011.
3 <http://linkinghub.elsevier.com/retrieve/pii/S2210784315000741>.
- 4 Chatterjee, N., P. K. R. Nair, S. Chakraborty, and V. D. Nair, 2018: Changes in soil carbon stocks
5 across the Forest-Agroforest-Agriculture/Pasture continuum in various agroecological regions:
6 A meta-analysis. *Agric. Ecosyst. Environ.*, **266**, 55–67.
- 7 Chen, Y., and Coauthors, 2018: Great uncertainties in modeling grazing impact on carbon
8 sequestration: a multi-model inter-comparison in temperate Eurasian Steppe. *Environ. Res. Lett.*,
9 **13**, 75005, doi:10.1088/1748-9326/aacc75. <http://stacks.iop.org/1748-9326/13/i=7/a=075005?key=crossref.8c1ea2666d647d78ac753b12554f5bb7>.
- 10
- 11 Chhetri, N., P. Chaudhary, P. R. Tiwari, and R. B. Yadaw, 2012: Institutional and technological
12 innovation: Understanding agricultural adaptation to climate change in Nepal. *Appl. Geogr.*, **33**,
13 142–150, doi:10.1016/J.APGEOG.2011.10.006.
14 <http://www.sciencedirect.com/science/article/pii/S0143622811001834> (Accessed January 11,
15 2018).
- 16 Chivenge, P., T. Mabhaudhi, A. T. Modi, and P. Mafongoya, 2015: The potential role of neglected
17 and underutilised crop species as future crops under water scarce conditions in Sub-Saharan
18 Africa. *Int. J. Environ. Res. Public Health*, **12**, 5685–5711, doi:10.3390/ijerph120605685.
- 19 Christmann, S., and Coauthors, 2009: *Food security and climate change in Central Asia and the*
20 *Caucasus*. Tashkent, Uzbekistan, 75 pp pp.
- 21 Chuwah, C., T. van Noije, D. P. van Vuuren, E. Stehfest, and W. Hazeleger, 2015: Global impacts of
22 surface ozone changes on crop yields and land use. *Atmos. Environ.*, **106**, 11–23,
23 doi:10.1016/j.atmosenv.2015.01.062. <http://dx.doi.org/10.1016/j.atmosenv.2015.01.062>.
- 24 CIPHET, 2007: *Vision - 2025 - CIPHET Perspective Plan, Central Institute of Post Harvest*
25 *Engineering and Technology*. P.O. PAU, LUDHIANA - 141 004, 74 pp.
26 https://www.ciphnet.in/upload/files/vision_2025.pdf.
- 27 Clapp, J., 2016: Food security and international trade--Unpacking disputed narratives. Background
28 paper prepared for The State of Agricultural Commodity Markets 2015--16. FAO: Rome.
- 29 Clay, E., 2006: Is food aid effective? *Id21 insights*, **61**, 6.
- 30 Cléménçon, R., 2016: The Two Sides of the Paris Climate Agreement : Dismal Failure or Historic
31 Breakthrough ? doi:10.1177/1070496516631362.
- 32 Clemens, M., J. Rijke, A. Pathirana, and J. Evers, 2015: Uncorrected Proof Uncorrected Proof. *J.*
33 *Water Clim. Chang.*, 1–13, doi:10.2166/wcc.2015.004.
- 34 De Clercq, D., Z. Wen, F. Fan, and L. Caicedo, 2016: Biomethane production potential from
35 restaurant food waste in megacities and project level-bottlenecks: A case study in Beijing.
36 *Renew. Sustain. Energy Rev.*, **59**, 1676–1685, doi:10.1016/J.RSER.2015.12.323.
37 <https://www.sciencedirect.com/science/article/pii/S1364032115017062> (Accessed May 30,
38 2018).
- 39 Clune, S., E. Crossin, and K. Verghese, 2017: Systematic review of greenhouse gas emissions for
40 different fresh food categories. *J. Clean. Prod.*, **140**, 766–783,
41 doi:10.1016/j.jclepro.2016.04.082.
- 42 Cohen, M. J., and J. L. Garrett, 2010: The food price crisis and urban food (in)security. *Environ.*
43 *Urban.*, **22**, 467–482, doi:10.1177/0956247810380375.
44 <http://journals.sagepub.com/doi/10.1177/0956247810380375> (Accessed November 3, 2018).
- 45 Cohn, A. S., A. Mosnier, P. Havlik, H. Valin, M. Herrero, E. Schmid, M. O’Hare, and M. Obersteiner,
46 2014: Cattle ranching intensification in Brazil can reduce global greenhouse gas emissions by
47 sparing land from deforestation. *Proc. Natl. Acad. Sci.*, **111**, 7236–7241,
48 doi:10.1073/pnas.1307163111.

- 1 Coleman-Jensen, A., M. P. Rabbitt, C. A. Gregory, and A. Singh, 2018: *Household Food Security in*
2 *the United States in 2017*, ERR-256. 44 pp pp. www.ers.usda.gov (Accessed October 8, 2018).
- 3 Čolović, D., S. Rakita, V. Banjac, O. Đuragić, and I. Čabarkapa, 2019: Plant Food By-Products as
4 Feed: Characteristics, Possibilities, Environmental Benefits, and Negative Sides. *Food Rev. Int.*,
5 1–27, doi:10.1080/87559129.2019.1573431.
- 6 Conant, R. T., 2010: *Challenges and opportunities for carbon sequestration in grassland systems A*
7 *technical report on grassland management and climate change mitigation*. Rome, 1-65 pp.
8 <https://pdfs.semanticscholar.org/9223/6427bd051d05cb2e5f5da13e3d68723c86cb.pdf> (Accessed
9 April 17, 2019).
- 10 Connell, J., 2015: Vulnerable islands: Climate change, tectonic change, and changing livelihoods in
11 the Western Pacific. *Contemp. Pac.*, **27**, 1–36, doi:10.1353/cp.2015.0014.
- 12 —, 2016: Last days in the Carteret Islands? Climate change, livelihoods and migration on coral
13 atolls. *Asia Pac. Viewp.*, **57**, 3–15, doi:10.1111/apv.12118.
14 <http://doi.wiley.com/10.1111/apv.12118> (Accessed April 16, 2019).
- 15 Conservation Evidence, 2019: Conservation Evidence: Synopses.
16 <https://www.conservationevidence.com/synopsis/index> (Accessed April 24, 2019).
- 17 Le Conte, Y., and M. Navajas, 2008: Climate change: impact on honey bee populations and diseases.
18 *Rev. Sci. Tech.*, **27**, 485–497, 499–510. <http://www.ncbi.nlm.nih.gov/pubmed/18819674>
19 (Accessed October 29, 2018).
- 20 Conway, G., 2013: *Sustainable Intensification: A New Paradigm for African Agriculture*. Montpellier
21 Panel.
- 22 Cook, B. I., K. J. Anchukaitis, R. Touchan, D. M. Meko, and E. R. Cook, 2016: Spatiotemporal
23 drought variability in the Mediterranean over the last 900 years. *J. Geophys. Res. Atmos.*, **121**,
24 2060–2074, doi:10.1002/2015JD023929. <http://doi.wiley.com/10.1002/2015JD023929>
25 (Accessed May 22, 2018).
- 26 Cook, S., L. Silici, B. Adolph, and S. Walker, 2015: *Sustainable intensification revisited*. IIED, 32 pp.
27 <http://pubs.iied.org/pdfs/14651IIED.pdf>.
- 28 Coomes, O., S. McGuire, E. Garine, S. Caillon, D. M.-F. Policy, and undefined 2015, Farmer seed
29 networks make a limited contribution to agriculture? Four common misconceptions. *Elsevier*.
- 30 Coomes, O. T., and Coauthors, 2015: Farmer seed networks make a limited contribution to
31 agriculture? Four common misconceptions. *Food Policy*, **56**, 41–50,
32 doi:10.1016/J.FOODPOL.2015.07.008.
- 33 Coumou, D., V. Petoukhov, S. Rahmstorf, S. Petri, and H. J. Schellnhuber, 2014: Quasi-resonant
34 circulation regimes and hemispheric synchronization of extreme weather in boreal summer.
35 *Proc. Natl. Acad. Sci.*, **111**, 12331–12336, doi:10.1073/pnas.1412797111.
36 <http://www.pnas.org/content/111/34/12331.abstract>.
- 37 Cramer, W., G. W. Yohe, M. Auffhammer, C. Huggel, U. Molau, M. A. F. da S. Dias, and R.
38 Leemans, 2014: Detection and attribution of observed impacts. *Climate Change 2014:*
39 *Impacts, Adaptation, and Vulnerability*, Cambridge University Press, 979–1038.
- 40 Craven, L. K., 2015: Migration-affected change and vulnerability in rural Vanuatu. *Asia Pac. Viewp.*,
41 **56**, 223–236, doi:10.1111/apv.12066.
- 42 Creutzig, F., B. Fernandez, H. Haberl, R. Khosla, Y. Mulugetta, and K. C. Seto, 2016: Beyond
43 Technology: Demand-Side Solutions for Climate Change Mitigation. *Annu. Rev. Environ.*
44 *Resour.*, **41**, 173–198, doi:10.1146/annurev-environ-110615-085428.
45 <http://www.annualreviews.org/doi/10.1146/annurev-environ-110615-085428>.
- 46 Crews, T. E., W. Carton, and L. Olsson, 2018: Is the future of agriculture perennial? Imperatives and
47 opportunities to reinvent agriculture by shifting from annual monocultures to perennial

- 1 polycultures. *Glob. Sustain.*, **1**, e11, doi:DOI: 10.1017/sus.2018.11.
2 [https://www.cambridge.org/core/article/is-the-future-of-agriculture-perennial-imperatives-and-](https://www.cambridge.org/core/article/is-the-future-of-agriculture-perennial-imperatives-and-opportunities-to-reinvent-agriculture-by-shifting-from-annual-monocultures-to-perennial-polycultures/0F69B1DBF3493462B4D46EB8F0F541EE)
3 [opportunities-to-reinvent-agriculture-by-shifting-from-annual-monocultures-to-perennial-](https://www.cambridge.org/core/article/is-the-future-of-agriculture-perennial-imperatives-and-opportunities-to-reinvent-agriculture-by-shifting-from-annual-monocultures-to-perennial-polycultures/0F69B1DBF3493462B4D46EB8F0F541EE)
4 [polycultures/0F69B1DBF3493462B4D46EB8F0F541EE](https://www.cambridge.org/core/article/is-the-future-of-agriculture-perennial-imperatives-and-opportunities-to-reinvent-agriculture-by-shifting-from-annual-monocultures-to-perennial-polycultures/0F69B1DBF3493462B4D46EB8F0F541EE).
- 5 Cruz-Blanco, M., C. Santos, P. Gavilán, and I. J. Lorite, 2015: Uncertainty in estimating reference
6 evapotranspiration using remotely sensed and forecasted weather data under the climatic
7 conditions of Southern Spain. *Int. J. Climatol.*, **35**, 3371–3384, doi:10.1002/joc.4215.
- 8 Curtis, P. G., C. M. Slay, N. L. Harris, A. Tyukavina, and M. C. Hansen, 2018: Classifying drivers of
9 global forest loss. *Science (80-.)*, **361**, 1108–1111, doi:10.1126/science.aau3445.
10 <http://science.sciencemag.org/content/361/6407/1108>.
- 11 Dagevos, H., and J. Voordouw, 2013: Sustainability and meat consumption: is reduction realistic?
12 *Sustain. Sci. Pract. Policy*, **9**, 60–69, doi:10.1080/15487733.2013.11908115.
13 <https://www.tandfonline.com/doi/full/10.1080/15487733.2013.11908115> (Accessed April 5,
14 2019).
- 15 Daioglou, V., J. C. Doelman, B. Wicke, A. Faaij, and D. P. van Vuuren, 2019: Integrated assessment
16 of biomass supply and demand in climate change mitigation scenarios. *Glob. Environ. Chang.*,
17 **54**, 88–101, doi:10.1016/J.GLOENVCHA.2018.11.012.
- 18 DaMatta, F. M., A. Grandis, B. C. Arenque, and M. S. Buckeridge, 2010: Impacts of climate changes
19 on crop physiology and food quality. *Food Res. Int.*, **43**, 1814–1823,
20 doi:10.1016/J.FOODRES.2009.11.001.
21 <https://www.sciencedirect.com/science/article/pii/S0963996909003421> (Accessed April 23,
22 2018).
- 23 Darmon, N., and A. Drewnowski, 2015: Contribution of food prices and diet cost to socioeconomic
24 disparities in diet quality and health: a systematic review and analysis. *Nutr. Rev.*, **73**, 643–660,
25 doi:10.1093/nutrit/nuv027. +.
- 26 Darnton-Hill, I., and B. Cogill, 2010: Maternal and Young Child Nutrition Adversely Affected by
27 External Shocks Such As Increasing Global Food Prices. *J. Nutr.*, **140**, 162S–169S,
28 doi:10.3945/jn.109.111682. <https://academic.oup.com/jn/article/140/1/162S/4600327> (Accessed
29 November 3, 2018).
- 30 Dasgupta, P., J. Morton, D. Dodman, B. Karapinar, F. Meza, M. Rivera-Ferre, A. Toure Sarr, and K.
31 Vincent, 2014: Rural Areas. *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part*
32 *A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment*
33 *Report of the Intergovernmental Panel on Climate Change*, C.B. Field et al., Eds., Cambridge
34 University Press, Cambridge, New York, 613–657 [http://www.ipcc.ch/pdf/assessment-](http://www.ipcc.ch/pdf/assessment-report/ar5/wg2/WGIIAR5-Chap9_FINAL.pdf)
35 [report/ar5/wg2/WGIIAR5-Chap9_FINAL.pdf](http://www.ipcc.ch/pdf/assessment-report/ar5/wg2/WGIIAR5-Chap9_FINAL.pdf) (Accessed December 8, 2017).
- 36 Davis, K. F., K. Yu, M. Herrero, P. Havlik, J. A. Carr, and P. D’Odorico, 2015: Historical trade-offs
37 of livestock’s environmental impacts. *Environ. Res. Lett.*, **10**, 125013.
- 38 Dawe, D., 2010: *The Rice Crisis: Markets, Policies and Food Security*. FAO and Earthscan, Rome,
39 Italy, 393 pp pp.
- 40 Deryng, D., D. Conway, N. Ramankutty, J. Price, and R. Warren, 2014: Global crop yield response to
41 extreme heat stress under multiple climate change futures. *Environ. Res. Lett.*, **9**, 34011,
42 doi:10.1088/1748-9326/9/3/034011. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/9/i=3/a=034011?key=crossref.dbe03166b591fb71af81975aae0cc1ee)
43 [9326/9/i=3/a=034011?key=crossref.dbe03166b591fb71af81975aae0cc1ee](http://stacks.iop.org/1748-9326/9/i=3/a=034011?key=crossref.dbe03166b591fb71af81975aae0cc1ee).
- 44 Descheemaeker, K., M. Zijlstra, P. Masikati, O. Crespo, and S. Homann-Kee Tui, 2018: Effects of
45 climate change and adaptation on the livestock component of mixed farming systems: A
46 modelling study from semi-arid Zimbabwe. *Agric. Syst.*, **159**, 282–295,
47 doi:10.1016/j.agsy.2017.05.004. <https://doi.org/10.1016/j.agsy.2017.05.004>.
- 48 Deutsch, C. A., J. J. Tewksbury, M. Tigchelaar, D. S. Battisti, S. C. Merrill, R. B. Huey, and R. L.
49 Naylor, 2018: Increase in crop losses to insect pests in a warming climate. *Science*, **361**, 916–

- 1 919, doi:10.1126/science.aat3466. <http://www.ncbi.nlm.nih.gov/pubmed/30166490> (Accessed
2 March 4, 2019).
- 3 Development Initiatives, 2017: *Global Nutrition Report 2017: Nourishing the SDGs*. Bristol, UK, 115
4 pp. https://www.globalnutritionreport.org/files/2017/11/Report_2017.pdf (Accessed January 11,
5 2018).
- 6 ———, 2018: *2018 Global Nutrition Report: Shining a light to spur action on nutrition*. Bristol, UK, 1-
7 161 pp. <http://ebrary.ifpri.org/cdm/ref/collection/p15738coll2/id/128484>.
- 8 Devereux, S., 2015: *Social Protection and Safety Nets in the Middle East and North Africa*.
9 www.ids.ac.uk/publications.
- 10 Dhanush, D., and S. Vermeulen, 2016: Climate change adaptation in agriculture: practices and
11 technologies Opportunities for climate action in agricultural systems. 1–7,
12 doi:10.13140/RG.2.1.4269.0321.
- 13 Dickie, I. A., and Coauthors, 2014: Conflicting values: ecosystem services and invasive tree
14 management. *Biol. Invasions*, **16**, 705–719, doi:10.1007/s10530-013-0609-6.
15 <http://link.springer.com/10.1007/s10530-013-0609-6> (Accessed April 25, 2019).
- 16 Din, K., S. Tariq, A. Mahmood, and G. Rasul, 2014: Temperature and Precipitation: GLOF
17 Triggering Indicators in Gilgit-Baltistan, Pakistan. *Pakistan J. Meteorol.*, **10**.
18 http://www.pmd.gov.pk/rnd/rndweb/rnd_new/journal/vol10_issue20_files/4.pdf (Accessed
19 January 9, 2018).
- 20 Dinesh, D., P. Aggarwal, A. Khatri-chhetri, A. Maria, L. Rodríguez, C. Mungai, L. Sebastian, and R.
21 Zougmore, 2017: The rise in Climate-Smart Agriculture strategies , policies , partnerships and
22 investments across the globe. *Agric. Dev.*, **30**, 4–9.
23 <https://cgspace.cgiar.org/handle/10568/81372>.
- 24 DKNC, D. K. N. C., 2012: Managing Climate Extremes and Disasters in the Agriculture Sector :
25 Lessons from the IPCC SREX Report Contents. *Clim. Dev. Knowl. Netw.*, 36.
26 [http://www.africaportal.org/dspace/articles/managing-climate-extremes-and-disasters-](http://www.africaportal.org/dspace/articles/managing-climate-extremes-and-disasters-agriculture-sector-lessons-ipcc-srex-report)
27 [agriculture-sector-lessons-ipcc-srex-report](http://www.africaportal.org/dspace/articles/managing-climate-extremes-and-disasters-agriculture-sector-lessons-ipcc-srex-report).
- 28 Doak, C. M., and B. M. Popkin, 2017: Overweight and Obesity. *Nutrition and Health in a Developing*
29 *World*, M.W. De Pee, Saskia, Taren, Douglas, Bloem, Ed., Springer, 143–158
30 <http://link.springer.com/10.1007/978-3-319-43739-2>.
- 31 Doelman, J. C., and Coauthors, 2018: Exploring SSP land-use dynamics using the IMAGE model:
32 Regional and gridded scenarios of land-use change and land-based climate change mitigation.
33 *Glob. Environ. Chang.*, **48**, 119–135, doi:<https://doi.org/10.1016/j.gloenvcha.2017.11.014>.
- 34 Donatti, C. I., C. A. Harvey, M. R. Martinez-Rodriguez, R. Vignola, and C. M. Rodriguez, 2018:
35 Vulnerability of smallholder farmers to climate change in Central America and Mexico: current
36 knowledge and research gaps. *Clim. Dev.*, 1–23, doi:10.1080/17565529.2018.1442796.
37 <https://www.tandfonline.com/doi/full/10.1080/17565529.2018.1442796> (Accessed April 16,
38 2019).
- 39 Dooley, K., and D. Stabinsky, 2018: *Missing Pathways to 1.5°C: The role of the land sector in*
40 *ambitious climate action*. climatelandambitionrightsalliance.org/report.
- 41 Van Dooren, C., M. Marinussen, H. Blonk, H. Aiking, and P. Vellinga, 2014: Exploring dietary
42 guidelines based on ecological and nutritional values: A comparison of six dietary patterns.
43 *Food Policy*, **44**, 36–46, doi:10.1016/j.foodpol.2013.11.002.
44 <http://dx.doi.org/10.1016/j.foodpol.2013.11.002>.
- 45 Doran-Browne, N. A., R. J. Eckard, R. Behrendt, and R. S. Kingwell, 2015: Nutrient density as a
46 metric for comparing greenhouse gas emissions from food production. *Clim. Change*, **129**, 73–
47 87, doi:10.1007/s10584-014-1316-8. <http://link.springer.com/10.1007/s10584-014-1316-8>
48 (Accessed April 5, 2019).

- 1 Dou, Y., R. Silva, H. Yang, and J. Lliu, 2018: Spillover effect offsets the conservation effort in the
2 Amazon. *Geogr. Sci.*, **28**, 1715–1732, doi:10.1007/s11442-018-1539-0.
- 3 Dowla, A., 2018: Climate change and microfinance. *Bus. Strateg. Dev.*, **1**, 78–87,
4 doi:10.1002/bsd2.13.
- 5 Dubbeling, M., and H. de Zeeuw, 2011: Urban Agriculture and Climate Change Adaptation: Ensuring
6 Food Security Through Adaptation BT - Resilient Cities. K. Otto-Zimmermann, Ed., Dordrecht,
7 Springer Netherlands, 441–449.
- 8 Dumont, A. M., G. Vanloqueren, P. M. Stassart, and P. V. Baret, 2016: Clarifying the socioeconomic
9 dimensions of agroecology: between principles and practices. *Agroecol. Sustain. Food Syst.*, **40**,
10 24–47, doi:10.1080/21683565.2015.1089967.
11 <http://www.tandfonline.com/doi/full/10.1080/21683565.2015.1089967> (Accessed March 4,
12 2019).
- 13 Dunne, J. P., R. J. Stouffer, and J. G. John, 2013: Reductions in labour capacity from heat stress under
14 climate warming. *Nat. Clim. Chang.*, **3**, 563, doi:10.1038/nclimate1827
15 <https://www.nature.com/articles/nclimate1827#supplementary-information>.
16 <http://dx.doi.org/10.1038/nclimate1827>.
- 17 Easter, T. S., A. K. Killion, and N. H. Carter, 2018: Climate change, cattle, and the challenge of
18 sustainability in a telecoupled system in Africa. *Ecol. Soc.*, **23**, art10, doi:10.5751/ES-09872-
19 230110. <https://www.ecologyandsociety.org/vol23/iss1/art10/> (Accessed April 17, 2019).
- 20 Edvardsson Björnberg, K., and S. O. Hansson, 2013: Gendering local climate adaptation. *Local*
21 *Environ.*, **18**, 217–232, doi:10.1080/13549839.2012.729571.
22 <http://www.tandfonline.com/doi/abs/10.1080/13549839.2012.729571> (Accessed November 3,
23 2018).
- 24 EFSA Panel on Biological Hazards, 2016: Growth of spoilage bacteria during storage and transport of
25 meat. *EFSA J.*, **14**, doi:10.2903/j.efsa.2016.4523.
- 26 Egeru, A., 2016: Climate risk management information, sources and responses in a pastoral region in
27 East Africa. *Clim. Risk Manag.*, **11**, 1–14, doi:10.1016/J.CRM.2015.12.001.
28 <https://www.sciencedirect.com/science/article/pii/S2212096315000418> (Accessed March 16,
29 2019).
- 30 El-Said, H., and J. Harrigan, 2014: Economic reform, social welfare, and instability: Jordan, Egypt,
31 Morocco, and Tunisia, 1983--2004. *Middle East J.*, **68**, 99–121.
- 32 Elum, Z. A., D. M. Modise, and A. Marr, 2017: Farmer's perception of climate change and responsive
33 strategies in three selected provinces of South Africa. *Clim. Risk Manag.*, **16**, 246–257,
34 doi:10.1016/j.crm.2016.11.001. <http://dx.doi.org/10.1016/j.crm.2016.11.001>.
- 35 Emberson, L. D., and Coauthors, 2018: Ozone effects on crops and consideration in crop models. *Eur.*
36 *J. Agron.*, 0–1, doi:10.1016/j.eja.2018.06.002. <https://doi.org/10.1016/j.eja.2018.06.002>.
- 37 End Hunger UK, 2018: Shocking figures showing hidden hunger show why we need to find out more
38 - End Hunger UK. [http://endhungeruk.org/shocking-figures-showing-hidden-hunger-show-need-
39 find/#more-274](http://endhungeruk.org/shocking-figures-showing-hidden-hunger-show-need-find/#more-274) (Accessed October 8, 2018).
- 40 Enenkel, M., and Coauthors, 2015: Drought and food security - Improving decision-support via new
41 technologies and innovative collaboration. *Glob. Food Sec.*, **4**, 51–55,
42 doi:10.1016/j.gfs.2014.08.005. <http://dx.doi.org/10.1016/j.gfs.2014.08.005>.
- 43 Englund, O., G. Berndes, and C. Cederberg, 2017: How to analyse ecosystem services in
44 landscapes—A systematic review. *Ecol. Indic.*, doi:10.1016/j.ecolind.2016.10.009.
- 45 Ensor, J. E., S. E. Park, E. T. Hoddy, and B. D. Ratner, 2015: A rights-based perspective on adaptive
46 capacity. *Glob. Environ. Chang.*, **31**, 38–49, doi:10.1016/J.GLOENVCHA.2014.12.005.
- 47 Ensor, J. E., S. E. Park, S. J. Attwood, A. M. Kaminski, and J. E. Johnson, 2018: Can community-

- 1 based adaptation increase resilience? *Clim. Dev.*, **10**, 134–151,
2 doi:10.1080/17565529.2016.1223595.
3 <https://www.tandfonline.com/doi/full/10.1080/17565529.2016.1223595> (Accessed October 23,
4 2018).
- 5 Erb, K.-H., T. Fetzl, T. Kastner, C. Kroisleitner, C. Lauk, A. Mayer, and M. Niedertscheider, 2016:
6 Livestock Grazing, the Neglected Land Use. *Social Ecology*, Springer International Publishing,
7 Cham, 295–313 http://link.springer.com/10.1007/978-3-319-33326-7_13 (Accessed April 5,
8 2019).
- 9 Erb, K. H., H. Haberl, M. R. Jepsen, T. Kuemmerle, M. Lindner, D. Müller, P. H. Verburg, and A.
10 Reenberg, 2013: A conceptual framework for analysing and measuring land-use intensity. *Curr.*
11 *Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2013.07.010.
- 12 Ercsey-Ravasz, M., Z. Toroczkai, Z. Lakner, and J. Baranyi, 2012: Complexity of the international
13 agro-food trade network and its impact on food safety. *PLoS One*, **7**, e37810.
- 14 European Union, 2018: Public-private startup accelerators in regional business support ecosystems.
15 13.
16 [https://www.interregeurope.eu/fileadmin/user_upload/plp_uploads/policy_briefs/Policy_Brief_-
17 _Public-private_startup_accelerators_in_regional_business_support_ecosystems_2018-12-
18 20_VALIDATED.pdf](https://www.interregeurope.eu/fileadmin/user_upload/plp_uploads/policy_briefs/Policy_Brief_-_Public-private_startup_accelerators_in_regional_business_support_ecosystems_2018-12-20_VALIDATED.pdf).
- 19 Eze, S. O., 2017: Constraints to climate change adaptation among cassava women farmers:
20 Implications for agricultural transformation and food security in ebonyi state, nigeria. *Int. J.*
21 *Ecosyst. Ecol. Sci.*, **7**, 219–228. <http://paper.researchbib.com/view/paper/103189> (Accessed
22 December 20, 2017).
- 23 Ezzine-de-Blas, D., S. Wunder, M. Ruiz-Pérez, and R. del P. Moreno-Sanchez, 2016: Global Patterns
24 in the Implementation of Payments for Environmental Services. *PLoS One*, **11**, e0149847,
25 doi:10.1371/journal.pone.0149847. <https://dx.plos.org/10.1371/journal.pone.0149847> (Accessed
26 November 5, 2018).
- 27 Fader, M., D. Gerten, M. Krause, W. Lucht, and W. Cramer, 2013: Spatial decoupling of agricultural
28 production and consumption: quantifying dependences of countries on food imports due to
29 domestic land and water constraints. *Environ. Res. Lett.*, **8**, 14046, doi:10.1088/1748-
30 9326/8/1/014046. [http://stacks.iop.org/1748-
31 9326/8/i=1/a=014046?key=crossref.7dd1e79e39bbf5d7423efaecb00e003](http://stacks.iop.org/1748-9326/8/i=1/a=014046?key=crossref.7dd1e79e39bbf5d7423efaecb00e003) (Accessed March 27,
32 2013).
- 33 Fagerli, R. A., and M. Wandel, 1999: Gender Differences in Opinions and Practices with Regard to a
34 “Healthy Diet.” *Appetite*, **32**, 171–190, doi:10.1006/APPE.1998.0188.
35 <https://www.sciencedirect.com/science/article/pii/S019566639890188X> (Accessed April 5,
36 2019).
- 37 Falkner, R., 2016: The Paris Agreement and the new logic of international climate politics. **5**, 1107–
38 1125.
- 39 Fanzo, J., R. McLaren, C. Davis, and J. Choufani, 2017: Climate change and variability: What are the
40 risks for nutrition, diets, and food systems? *IFPRI*.
- 41 FAO, 2001a: *Food Insecurity in the World 2001*. <http://www.fao.org/3/a-y1500e.pdf> (Accessed
42 December 6, 2017).
- 43 —, 2001b: *Food Balance Sheets: A Handbook (Rome: FAO)*. Rome, Italy, 99 pp.
44 <http://www.fao.org/3/a-x9892e.pdf>.
- 45 —, 2005: *Voluntary Guidelines to support the progressive realization of the right to adequate food*
46 *in the context of the national food security*. The Right to Food. Food and Agriculture
47 Organization of the United Nations, Rome, Italy, 48 pp. [http://www.fao.org/right-to-
48 food/resources/resources-detail/en/c/44965/](http://www.fao.org/right-to-food/resources/resources-detail/en/c/44965/) (Accessed April 14, 2019).

- 1 —, 2007: *The State of Food and Agriculture 2007*. Rome, Italy, 222p pp.
- 2 —, 2008: *Food security in mountains – High time for action*. Rome, Italy,
3 http://www.fao.org/fileadmin/templates/mountainday/docs/pdf_2008/IMD08_leaflet_En_LR.pdf
4 .
- 5 —, 2011a: *The State of Food and Agriculture. Women in Agriculture: Closing the gender gap for*
6 *development*. Rome, 160 pp. <http://www.fao.org/docrep/013/i2050e/i2050e.pdf> (Accessed
7 December 8, 2017).
- 8 —, 2011b: *Global food losses and food waste: extent, causes and prevention*. FAO, Rome,.
- 9 —, 2011c: *Dietary protein quality evaluation in human nutrition*. 1-79 pp.
10 <http://www.fao.org/ag/humannutrition/35978-02317b979a686a57aa4593304ffc17f06.pdf>.
- 11 —, 2011d: *Save and Grow. A Policymaker's Guide to the Sustainable Intensification of*
12 *Smallholder Crop Production*. Rome, 1-116 pp. [http://www.fao.org/policy-](http://www.fao.org/policy-support/resources/resources-details/en/c/421716/)
13 [support/resources/resources-details/en/c/421716/](http://www.fao.org/policy-support/resources/resources-details/en/c/421716/).
- 14 —, 2011e: *Save and Grow. A Policymaker's Guide to the Sustainable Intensification of*
15 *Smallholder Crop Production*. Rome, 1-116 pp.
- 16 —, 2012: *Incorporating climate change considerations into agricultural investment*. 148 pp.
- 17 —, 2013a: *The State of Food and Agriculture*. Rome, 114 pp.
18 <http://www.fao.org/docrep/018/i3300e/i3300e.pdf> (Accessed April 20, 2018).
- 19 —, 2013b: *Food wastage footprint. Impacts on natural resources*. 63 pp.
20 <http://www.fao.org/docrep/018/i3347e/i3347e.pdf>.
- 21 —, 2013c: *Climate-Smart Agriculture: Managing Ecosystems for Sustainable Livelihoods*. 1-14 pp.
22 <http://www.fao.org/3/a-an177e.pdf> (Accessed April 24, 2019).
- 23 —, 2014a: *The State of Food Insecurity in the World (SOFI) 2014*. Rome, Italy,.
- 24 —, 2014b: *Food Wastage Footprint: Full-cost Accounting*. 98 pp. [http://www.fao.org/3/a-](http://www.fao.org/3/a-i3991e.pdf)
25 [i3991e.pdf](http://www.fao.org/3/a-i3991e.pdf).
- 26 —, 2015a: *Food Wastage Footprint and Climate Change*. Rome, 4 pp.
27 [http://www.fao.org/fileadmin/templates/nr/sustainability_pathways/docs/FWF_and_climate_](http://www.fao.org/fileadmin/templates/nr/sustainability_pathways/docs/FWF_and_climate_change.pdf)
28 [nge.pdf](http://www.fao.org/fileadmin/templates/nr/sustainability_pathways/docs/FWF_and_climate_change.pdf) (Accessed May 24, 2018).
- 29 —, 2015b: *L'Insécurité Alimentaire Dans Le Monde*.
- 30 —, 2015c: *Global trends in GDP and Agriculture Value Added (1970-2013)*. Rome, Italy, 1-6 pp.
31 [http://www.fao.org/fileadmin/templates/ess/documents/GDP/IND_NewsRelease_EN_27Apr20](http://www.fao.org/fileadmin/templates/ess/documents/GDP/IND_NewsRelease_EN_27Apr2015_.pdf)
32 [15_.pdf](http://www.fao.org/fileadmin/templates/ess/documents/GDP/IND_NewsRelease_EN_27Apr2015_.pdf).
- 33 —, 2016a: *2015–2016 El Niño - Early action and response for agriculture, food security and*
34 *nutrition - UPDATE #10*. Rome (Italy), 43 pp. <http://www.fao.org/3/a-i6049e.pdf> (Accessed
35 March 9, 2019).
- 36 —, 2016b: *State of the World's forests*. Rome, Italy, 126 pp pp. <http://www.fao.org/3/a-i5588e.pdf>.
- 37 —, 2016c: *Farmer Field School: Guidance Document Planning for Quality Programmes*. Food and
38 Agriculture Organization of the United Nations, Ed. Food and Agriculture Organization of the
39 United Nations, [https://fr.scribd.com/book/313692418/Farmer-Field-School-Guidance-](https://fr.scribd.com/book/313692418/Farmer-Field-School-Guidance-Documents-Planning-for-Quality-Programmes)
40 [Document-Planning-for-Quality-Programmes](https://fr.scribd.com/book/313692418/Farmer-Field-School-Guidance-Documents-Planning-for-Quality-Programmes) (Accessed April 14, 2019).
- 41 —, 2017: *Livestock solutions for climate change*. Rome, 1-8 pp. <http://www.fao.org/3/a-i8098e.pdf>
42 (Accessed April 17, 2019).
- 43 —, 2018a: *The future of food and agriculture: Alternative pathways to 2050*. 228 pp.
44 <http://www.fao.org/publications/fofa/en/>.
- 45 —, 2018b: *The future of food and agriculture: Trends and challenges*. Rome, 180 pp.

- 1 <http://www.fao.org/3/a-i6583e.pdf> (Accessed May 18, 2018).
- 2 —, 2018c: *The state of world fisheries and aquaculture 2018 : meeting the sustainable development*
3 *goals*. 210 pp. <http://www.fao.org/documents/card/en/c/I9540EN> (Accessed April 13, 2019).
- 4 —, 2018d: *The State of Agricultural Commodity Markets 2018*. Rome, Italy, 112 pp.
5 <http://www.fao.org/3/I9542EN/i9542en.pdf>.
- 6 —, 2018e: *The FAO Strategic Objectives*. Rome, Italy, <http://www.fao.org/3/a-mg994e.pdf>.
- 7 —, 2018f: *Gender and food loss in sustainable food value chains: A guiding note*. Rome, 1-56 pp.
8 <http://www.fao.org/3/I8620EN/i8620en.pdf> (Accessed April 17, 2019).
- 9 —, 2018g: *Globally Important Agricultural Heritage Systems (GIAHS). Combining agricultural*
10 *biodiversity, resilient ecosystems, traditional farming practices and cultural identity*. Rome, 1-
11 48 pp. <http://www.fao.org/3/i9187en/I9187EN.pdf> (Accessed April 17, 2019).
- 12 —, 2019a: Context for Sustainable Intensification of Agriculture. *Sustain. Food Agric.*, 171–172,
13 doi:10.1016/B978-0-12-812134-4.00010-8.
14 <https://www.sciencedirect.com/science/article/pii/B9780128121344000108> (Accessed April 16,
15 2019).
- 16 —, 2019b: SAVE FOOD: Global Initiative on Food Loss and Waste Reduction | Food and
17 Agriculture Organization of the United Nations. <http://www.fao.org/save-food/en/> (Accessed
18 April 24, 2019).
- 19 —, Mountain Partnership, UNCCD, SDC, and CDE, 2011: *Highlands and Drylands Mountains, a*
20 *source of resilience in arid regions*. Food and Agriculture Organization of FAO, UNCCD,
21 Mountain Partnership, Swiss Agency for Development and Cooperation, and CDE, Rome, 115
22 pp. <https://boris.unibe.ch/6754/1/i2248e00.pdf>.
- 23 —, IFAD, UNICEF, WFP, and WHO, 2017: *The State of Food Security and Nutrition in the World*
24 *2017. Building resilience for peace and food security*. Rome, Italy, 132 pp.
25 <http://www.fao.org/3/a-I7695e.pdf>.
- 26 —, —, —, —, and —, 2018: *The State of Food Security and Nutrition in the World 2018.*
27 *Building climate resilience for food security and nutrition*. Rome, www.fao.org/publications
28 (Accessed October 1, 2018).
- 29 FAOSTAT, 2018: Food and Agriculture Organization Corporate Statistical Database (FAOSTAT).
30 <http://www.fao.org/faostat/en/#home> (Accessed December 3, 2017).
- 31 Farbotko, C., 2010: Wishful sinking: Disappearing islands, climate refugees and cosmopolitan
32 experimentation. *Asia Pac. Viewp.*, **51**, 47–60, doi:10.1111/j.1467-8373.2010.001413.x.
33 <http://doi.wiley.com/10.1111/j.1467-8373.2010.001413.x> (Accessed May 8, 2018).
- 34 —, and H. Lazrus, 2012: The first climate refugees? Contesting global narratives of climate change
35 in Tuvalu. *Glob. Environ. Chang.*, **22**, 382–390, doi:10.1016/J.GLOENVCHA.2011.11.014.
36 <https://www.sciencedirect.com/science/article/pii/S0959378011002019?via%3Dihub> (Accessed
37 May 8, 2018).
- 38 Farchi, S., M. De Sario, E. Lapucci, M. Davoli, and P. Michelozzi, 2017: Meat consumption reduction
39 in Italian regions: Health co-benefits and decreases in GHG emissions. *PLoS One*, **12**, 1–19,
40 doi:10.1371/journal.pone.0182960. <http://dx.doi.org/10.1371/journal.pone.0182960>.
- 41 Farnworth, C. R., F. Baudron, J. A. Andersson, M. Misiko, L. Badstue, and C. M. Stirling, 2016:
42 Gender and conservation agriculture in East and Southern Africa: towards a research agenda.
43 *Int. J. Agric. Sustain.*, **14**, 142–165, doi:10.1080/14735903.2015.1065602.
44 <https://www.tandfonline.com/doi/full/10.1080/14735903.2015.1065602> (Accessed October 29,
45 2018).
- 46 Faye, B., and Coauthors, 2018a: Impacts of 1.5 versus 2.0 °c on cereal yields in the West African
47 Sudan Savanna. *Environ. Res. Lett.*, **13**, doi:10.1088/1748-9326/aaab40.

- 1 ———, H. Webber, M. Diop, M. L. Mbaye, J. D. Owusu-Sekyere, J. B. Naab, and T. Gaiser, 2018b:
2 Potential impact of climate change on peanut yield in Senegal, West Africa. *F. Crop. Res.*, **219**,
3 148–159, doi:10.1016/j.fcr.2018.01.034. <https://doi.org/10.1016/j.fcr.2018.01.034>.
- 4 Fearnside, P. M., 2015: Environment: Deforestation soars in the Amazon. *Nat.* 2015 5217553,.
- 5 van der Fels-Klerx, H. J., J. E. Olesen, M. S. Madsen, and P. W. Goedhart, 2012a: Climate change
6 increases deoxynivalenol contamination of wheat in north-western Europe. *Food Addit. Contam.*
7 *Part a-Chemistry Anal. Control Expo. Risk Assess.*, **29**, 1593–1604,
8 doi:10.1080/19440049.2012.691555.
- 9 ———, ———, L. J. Naustvoll, Y. Friocourt, M. J. B. Mengelers, and J. H. Christensen, 2012b: Climate
10 change impacts on natural toxins in food production systems, exemplified by deoxynivalenol in
11 wheat and diarrhetic shellfish toxins. *Food Addit. Contam. Part a-Chemistry Anal. Control*
12 *Expo. Risk Assess.*, **29**, 1647–1659, doi:10.1080/19440049.2012.714080.
- 13 Van der Fels-Klerx, H. J., E. D. van Asselt, M. S. Madsen, and J. E. Olesen, 2013: Impact of Climate
14 Change Effects on Contamination of Cereal Grains with Deoxynivalenol. *PLoS One*, **8**,
15 doi:10.1371/journal.pone.0073602.
- 16 Feng, S., A. B. Krueger, and M. Oppenheimer, 2010: Linkages among climate change, crop yields
17 and Mexico--US cross-border migration. *Proc. Natl. Acad. Sci.*, **107**, 14257–14262.
- 18 Feng, Z., and Coauthors, 2015: Constraints to nitrogen acquisition of terrestrial plants under elevated
19 CO₂. *Glob. Chang. Biol.*, **21**, 3152–3168, doi:10.1111/gcb.12938.
- 20 Fenghua, S., Y. Xiuqun, L. Shuang, and others, 2006: The contrast analysis on the average and
21 extremum temperature trend in Northeast China. *Sci. Meteorol. Sin.*, **26**, 157–163.
- 22 Fenn, B., 2018: *Impacts of CASH on Nutrition Outcomes: From Available Scientific Evidence to*
23 *Informed Actions.* 1-24 pp. [http://www.cashlearning.org/downloads/user-submitted-](http://www.cashlearning.org/downloads/user-submitted-resources/2018/06/1529400438.WFP-0000071735.pdf)
24 [resources/2018/06/1529400438.WFP-0000071735.pdf](http://www.cashlearning.org/downloads/user-submitted-resources/2018/06/1529400438.WFP-0000071735.pdf).
- 25 Fenton, A., and J. Paavola, 2015: Microfinance and climate change adaptation : an overview of the
26 current literature. **26**.
- 27 Ferrière, N., and A. Suwa-Eisenmann, 2015: Does Food Aid Disrupt Local Food Market? Evidence
28 from Rural Ethiopia. *World Dev.*, **76**, 114–131, doi:10.1016/J.WORLDDEV.2015.07.002.
29 <https://www.sciencedirect.com/science/article/pii/S0305750X15001618> (Accessed October 30,
30 2018).
- 31 Fetzel, T., and Coauthors, 2017: Quantification of uncertainties in global grazing systems assessment.
32 *Global Biogeochem. Cycles*, **31**, 1089–1102, doi:10.1002/2016GB005601.
33 <http://doi.wiley.com/10.1002/2016GB005601> (Accessed November 4, 2018).
- 34 Finer, M., S. Novoa, M. J. Weisse, R. Petersen, J. Mascaro, T. Souto, F. Stearns, and R. G. Martinez,
35 2018: 1Combating deforestation: From satellite to intervention. *Science (80-.)*, **360**, 1303–
36 1305, doi:10.1126/science.aat1203.
37 <http://www.sciencemag.org/lookup/doi/10.1126/science.aat1203>.
- 38 Finn, J. A., and Coauthors, 2013: Ecosystem function enhanced by combining four functional types of
39 plant species in intensively managed grassland mixtures: a 3-year continental-scale field
40 experiment. *J. Appl. Ecol.*, **50**, 365–375, doi:10.1111/1365-2664.12041.
- 41 Fishedick, M., and Coauthors, 2014: Industry. *Climate Change 2014: Mitigation of Climate Change.*
42 *Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental*
43 *Panel on Climate Change*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge,
44 United Kingdom and New York, NY, USA.
- 45 Fischer, J., and Coauthors, 2014: Land Sparing Versus Land Sharing: Moving Forward. *Conserv.*
46 *Let.*, **7**, 149–157, doi:10.1111/conl.12084.
- 47 FishStat, 2019: FishStatJ - Software for Fishery and Aquaculture Statistical Time Series.

- 1 <http://www.fao.org/fishery/statistics/software/fishstatj/en>.
- 2 Fitton, N., and Coauthors, 2017: Modelling spatial and inter-annual variations of nitrous oxide
3 emissions from UK cropland and grasslands using DailyDayCent. *Agric. Ecosyst. Environ.*, **250**,
4 1–11, doi:10.1016/J.AGEE.2017.08.032.
5 <https://www.sciencedirect.com/science/article/pii/S0167880917303894> (Accessed May 25,
6 2018).
- 7 Fjelde, H., and N. von Uexkull, 2012: Climate triggers: Rainfall anomalies, vulnerability and
8 communal conflict in Sub-Saharan Africa. *Polit. Geogr.*, **31**, 444–453,
9 doi:10.1016/J.POLGEO.2012.08.004.
10 <https://www.sciencedirect.com/science/article/pii/S0962629812001072> (Accessed March 15,
11 2019).
- 12 Folami, O. M., and A. O. Folami, 2013: Climate Change and Inter-Ethnic Conflict in Nigeria. *Peace*
13 *Rev.*, **25**, 104–110, doi:10.1080/10402659.2013.759783.
- 14 Ford, J. D., L. Berrang-Ford, A. Bunce, C. McKay, M. Irwin, and T. Pearce, 2015: The status of
15 climate change adaptation in Africa and Asia. *Reg. Environ. Chang.*, **15**, 801–814,
16 doi:10.1007/s10113-014-0648-2. <http://dx.doi.org/10.1007/s10113-014-0648-2>.
- 17 Forsyth, T., 2013: Community-based adaptation: a review of past and future challenges. *Wiley*
18 *Interdiscip. Rev. Clim. Chang.*, **4**, 439–446, doi:10.1002/wcc.231.
19 <http://doi.wiley.com/10.1002/wcc.231> (Accessed October 23, 2018).
- 20 Francis, C., 2016: The carbon farming solution: a global toolkit of perennial crops and regenerative
21 agriculture practices for climate change mitigation and food security, by Eric Toensmeier.
22 *Agroecol. Sustain. Food Syst.*, **40**, 1039–1040, doi:10.1080/21683565.2016.1214861.
23 <https://doi.org/10.1080/21683565.2016.1214861>.
- 24 Francis, C., and Coauthors, 2003: Agroecology: The Ecology of Food Systems. *J. Sustain. Agric.*, **22**,
25 99–118, doi:10.1300/J064v22n03_10.
26 http://www.tandfonline.com/doi/abs/10.1300/J064v22n03_10 (Accessed March 3, 2019).
- 27 Frank, S., and Coauthors, 2017: Reducing greenhouse gas emissions in agriculture without
28 compromising food security? *Environ. Res. Lett.*, **12**, 105004, doi:10.1088/1748-9326/aa8c83.
- 29 Frank, S., and Coauthors, 2018: Structural change as a key component for agricultural non-CO2
30 mitigation efforts. *Nat. Commun.*, **9**, 1060, doi:10.1038/s41467-018-03489-1.
31 <http://www.nature.com/articles/s41467-018-03489-1>.
- 32 Franzaring, J., I. Holz, and A. Fangmeier, 2013: Responses of old and modern cereals to CO2-
33 fertilisation. *Crop Pasture Sci.*, **64**, 943–956, doi:10.1071/cp13311.
- 34 Fratkin, E., E. A. Roth, and M. A. Nathan, 2004: Pastoral Sedentarization and Its Effects on
35 Children's Diet, Health, and Growth Among Rendille of Northern Kenya. *Hum. Ecol.*, **32**, 531–
36 559, doi:10.1007/s10745-004-6096-8. <http://link.springer.com/10.1007/s10745-004-6096-8>
37 (Accessed March 24, 2019).
- 38 Friel, S., and Coauthors, 2009: Public health benefits of strategies to reduce greenhouse-gas
39 emissions: food and agriculture. *Lancet*, **374**, 2016–2025, doi:10.1016/S0140-6736(09)61753-0.
40 [http://dx.doi.org/10.1016/S0140-6736\(09\)61753-0](http://dx.doi.org/10.1016/S0140-6736(09)61753-0).
- 41 Friis, C., J. Ø. Nielsen, I. Otero, H. Haberl, J. Niewöhner, and P. Hostert, 2016: From teleconnection
42 to telecoupling: taking stock of an emerging framework in land system science. *J. Land Use Sci.*,
43 **11**, 131–153, doi:10.1080/1747423X.2015.1096423.
44 <http://www.tandfonline.com/doi/full/10.1080/1747423X.2015.1096423> (Accessed April 17,
45 2019).
- 46 Fröhlich, C., and S. Klepp, 2019: Effects of Climate Change on Migration Crises in Oceania. *The*
47 *Oxford Handbook of Migration Crises*, C. Menjívar, M. Ruiz, and I. Ness, Eds., Oxford
48 University Press, 330–346

- 1 <http://oxfordhandbooks.com/view/10.1093/oxfordhb/9780190856908.001.0001/oxfordhb->
2 [9780190856908-e-52](http://oxfordhandbooks.com/view/10.1093/oxfordhb/9780190856908.001.0001/oxfordhb-9780190856908-e-52) (Accessed April 1, 2019).
- 3 Fujimori, S., T. Hasegawa, T. Masui, K. Takahashi, D. S. Herran, H. Dai, Y. Hijioka, and M.
4 Kainuma, 2017: SSP3: AIM implementation of Shared Socioeconomic Pathways. *Glob.*
5 *Environ. Chang.*, **42**, 268–283, doi:10.1016/j.gloenvcha.2016.06.009.
6 <http://dx.doi.org/10.1016/j.gloenvcha.2016.06.009>.
- 7 ———, ———, J. Rogelj, X. Su, P. Havlik, V. Krey, K. Takahashi, and K. Riahi, 2018: Inclusive climate
8 change mitigation and food security policy under 1.5 °C climate goal. *Environ. Res. Lett.*, **13**,
9 74033, doi:10.1088/1748-9326/aad0f7. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997)
10 [9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997).
- 11 Fujisaki, K., T. Chevallier, L. Chapuis-Lardy, A. Albrecht, T. Razafimbelo, D. Masse, Y. B. Ndour,
12 and J. L. Chotte, 2018: Soil carbon stock changes in tropical croplands are mainly driven by
13 carbon inputs: A synthesis. *Agric. Ecosyst. Environ.*, **259**, 147–158,
14 doi:10.1016/j.agee.2017.12.008.
- 15 Fuller, T., and M. Qingwen, 2013: Understanding Agricultural Heritage Sites as Complex Adaptive
16 Systems: The Challenge of Complexity. *J. Resour. Ecol.*, **4**, 195–201, doi:10.5814/j.issn.1674-
17 764x.2013.03.002. [http://www.bioone.org/doi/abs/10.5814/j.issn.1674-](http://www.bioone.org/doi/abs/10.5814/j.issn.1674-764x.2013.03.002)
18 [764x.2013.03.002](http://www.bioone.org/doi/abs/10.5814/j.issn.1674-764x.2013.03.002). (Accessed November 3, 2018).
- 19 Fulton, S., and Coauthors, 2019: From Fishing Fish to Fishing Data: The Role of Artisanal Fishers in
20 Conservation and Resource Management in Mexico. *Viability and Sustainability of Small-Scale*
21 *Fisheries in Latin America and The Caribbean*, S. Salas, M.J. Barragán-Paladines, and R.
22 Chuenpagdee, Eds., Springer, Cham, 151–175 [http://link.springer.com/10.1007/978-3-319-](http://link.springer.com/10.1007/978-3-319-76078-0_7)
23 [76078-0_7](http://link.springer.com/10.1007/978-3-319-76078-0_7) (Accessed November 2, 2018).
- 24 Fuss, S., and Coauthors, 2016: Research priorities for negative emissions. *Environ. Res. Lett.*, **11**,
25 115007, doi:10.1088/1748-9326/11/11/115007. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=11/a=115007?key=crossref.0d8b5da77c87ad502a8129cad2731d44)
26 [9326/11/i=11/a=115007?key=crossref.0d8b5da77c87ad502a8129cad2731d44](http://stacks.iop.org/1748-9326/11/i=11/a=115007?key=crossref.0d8b5da77c87ad502a8129cad2731d44).
- 27 Fuss, S., and Coauthors, 2018: Negative emissions - Part 2: Costs, potentials and side effects.
28 *Environ. Res. Lett.*, **13**, 63002, doi:10.1088/1748-9326/aabf9f. [https://doi.org/10.1088%2F1748-](https://doi.org/10.1088%2F1748-9326%2Faabf9f)
29 [9326%2Faabf9f](https://doi.org/10.1088%2F1748-9326%2Faabf9f).
- 30 Gabrielsson, S., and V. Ramasar, 2013: Widows: agents of change in a climate of water uncertainty. *J.*
31 *Clean. Prod.*, **60**, 34–42, doi:10.1016/J.JCLEPRO.2012.01.034.
32 <https://www.sciencedirect.com/science/article/pii/S0959652612000546?via%3Dihub> (Accessed
33 October 29, 2018).
- 34 Van Gaelen, H., E. Vanuytrecht, P. Willems, J. Diels, and D. Raes, 2017: Bridging rigorous
35 assessment of water availability from field to catchment scale with a parsimonious agro-
36 hydrological model. *Environ. Model. Softw.*, **94**, 140–156, doi:10.1016/j.envsoft.2017.02.014.
37 <http://dx.doi.org/10.1016/j.envsoft.2017.02.014>.
- 38 Gale, P., T. Drew, L. P. Phipps, G. David, and M. Wooldridge, 2009: The effect of climate change on
39 the occurrence and prevalence of livestock diseases in Great Britain: a review. *J. Appl.*
40 *Microbiol.*, **106**, 1409–1423, doi:10.1111/j.1365-2672.2008.04036.x.
- 41 Gamba, R. J., J. Schuchter, C. Rutt, and E. Y. W. Seto, 2015: Measuring the Food Environment and
42 its Effects on Obesity in the United States: A Systematic Review of Methods and Results. *J.*
43 *Community Health*, **40**, 464–475, doi:10.1007/s10900-014-9958-z.
44 <http://link.springer.com/10.1007/s10900-014-9958-z> (Accessed October 8, 2018).
- 45 Gao, C., 2018: The future of CRISPR technologies in agriculture. *Nat. Rev. Mol. Cell Biol.*, **19**, 275–
46 276, doi:10.1038/nrm.2018.2. <http://www.nature.com/doi/10.1038/nrm.2018.2> (Accessed
47 April 17, 2019).
- 48 Garcia Collazo, M. A., A. Panizza, and J. M. Paruelo, 2013: Ordenamiento Territorial de Bosques
49 Nativos: Resultados de la Zonificación realizada por provincias del Norte. *Ecol. Aust.*, **23**, 97–

- 1 107.
- 2 Garibaldi, L. A., and Coauthors, 2016: Mutually beneficial pollinator diversity and crop yield
3 outcomes in small and large farms. *Science* (80-.), **351**, 388–391,
4 doi:10.1126/SCIENCE.AAC7287. <https://science.sciencemag.org/content/351/6271/388>
5 (Accessed April 25, 2019).
- 6 —, B. Gemmill-Herren, R. D’Annolfo, B. E. Graeub, S. A. Cunningham, and T. D. Breeze, 2017:
7 Farming Approaches for Greater Biodiversity, Livelihoods, and Food Security. *Trends Ecol.*
8 *Evol.*, **32**, 68–80, doi:10.1016/j.tree.2016.10.001. <http://dx.doi.org/10.1016/j.tree.2016.10.001>.
- 9 Garnett, T., 2011: Where are the best opportunities for reducing greenhouse gas emissions in the food
10 system (including the food chain)? *Food Policy*, **36**, S23–S32,
11 doi:10.1016/J.FOODPOL.2010.10.010.
12 <https://www.sciencedirect.com/science/article/pii/S0306919210001132> (Accessed April 16,
13 2019).
- 14 —, 2016: Plating up solutions. *Science* (80-.), **353**, 1202–1204, doi:10.1126/science.aah4765.
- 15 Garnett, T., and Coauthors, 2013: Sustainable Intensification in Agriculture: Premises and Policies.
16 *Science* (80-.), **341**, 33–34, doi:10.1126/science.1234485.
17 <http://www.sciencemag.org/cgi/doi/10.1126/science.1234485>.
- 18 Garnett, T., and Coauthors, 2017: *Grazed and confused?* 127 pp.
- 19 Gasparri, N. I., and Y. le P. de Waroux, 2015: The Coupling of South American Soybean and Cattle
20 Production Frontiers: New Challenges for Conservation Policy and Land Change Science.
21 *Conserv. Lett.*, **8**, 290–298, doi:10.1111/conl.12121. <http://doi.wiley.com/10.1111/conl.12121>.
- 22 Gasper, R., A. Blohm, and M. Ruth, 2011: Social and economic impacts of climate change on the
23 urban environment. *Curr. Opin. Environ. Sustain.*, **3**, 150–157,
24 doi:10.1016/J.COSUST.2010.12.009.
25 <https://www.sciencedirect.com/science/article/pii/S187734351000148X> (Accessed May 25,
26 2018).
- 27 Gephart, J. A., L. Deutsch, M. L. Pace, M. Troell, and D. A. Seekell, 2017: Shocks to fish production:
28 Identification, trends, and consequences. *Glob. Environ. Chang.*, **42**, 24–32,
29 doi:10.1016/j.gloenvcha.2016.11.003.
- 30 Gerber, J. S., and Coauthors, 2016: Spatially explicit estimates of N2O emissions from croplands
31 suggest climate mitigation opportunities from improved fertilizer management. *Glob. Chang.*
32 *Biol.*, **22**, 3383–3394, doi:10.1111/gcb.13341.
- 33 Gerber, P. J., H. Steinfeld, B. Henderson, and others, 2013: *Tackling climate change through*
34 *livestock: a global assessment of emissions and mitigation opportunities*. Food and Agriculture
35 Organization of the United Nations, Rome,.
- 36 Gerber, P. J., A. Mottet, C. I. Opio, A. Falcucci, and F. Teillard, 2015: Environmental impacts of beef
37 production: Review of challenges and perspectives for durability. *Meat Sci.*, **109**, 2–12,
38 doi:10.1016/J.MEATSCI.2015.05.013.
- 39 Ghosh-Jerath, S., A. Singh, P. Kamboj, G. Goldberg, and M. S. Magsumbol, 2015: Traditional
40 Knowledge and Nutritive Value of Indigenous Foods in the Oraon Tribal Community of
41 Jharkhand: An Exploratory Cross-sectional Study. *Ecol. Food Nutr.*, **54**, 493–519,
42 doi:10.1080/03670244.2015.1017758.
43 <http://www.tandfonline.com/doi/full/10.1080/03670244.2015.1017758> (Accessed October 21,
44 2018).
- 45 Ghude, S. D., C. Jena, D. M. Chate, G. Beig, G. G. Pfister, R. Kumar, and V. Ramanathan, 2014:
46 Reductions in India’s crop yield due to ozone. *Geophys. Res. Lett.*, **41**, 5685–5691,
47 doi:10.1002/2014GL060930. <http://doi.wiley.com/10.1002/2014GL060930> (Accessed April 10,
48 2018).

- 1 Gibbs, H. K., and J. M. Salmon, 2015: Mapping the world's degraded lands. *Appl. Geogr.*, **57**, 12–21,
2 doi:10.1016/J.APGEOG.2014.11.024.
- 3 Gibbs, H. K., and Coauthors, 2015: Brazil's Soy Moratorium. *Science (80-.)*, **347**, 377–378,
4 doi:10.1126/science.aaa0181. <http://science.sciencemag.org/content/347/6220/377>.
- 5 Gil, J., M. Siebold, and T. Berger, 2015: Adoption and development of integrated crop-livestock-
6 forestry systems in Mato Grosso, Brazil. *Agric. Ecosyst. Environ.*, **199**, 394–406,
7 doi:10.1016/j.agee.2014.10.008.
- 8 Gil, J. D. B., and Coauthors, 2018: Tradeoffs in the quest for climate smart agricultural intensification
9 in Mato Grosso, Brazil. *Environ. Res. Lett.*, **13**, 64025, doi:10.1088/1748-9326/aac4d1.
- 10 Giller, K. E., E. Witter, M. Corbeels, and P. Tittonell, 2009: Conservation agriculture and smallholder
11 farming in Africa: The heretics' view. *F. Crop. Res.*, **114**, 23–34, doi:10.1016/j.fcr.2009.06.017.
- 12 Gliessman, S., 2018: Defining Agroecology. *Agroecol. Sustain. Food Syst.*, **42**, 599–600,
13 doi:10.1080/21683565.2018.1432329.
14 <https://www.tandfonline.com/doi/full/10.1080/21683565.2018.1432329> (Accessed April 24,
15 2019).
- 16 Gliessman, S. R., and E. Engles, 2014: *Agroecology: the ecology of sustainable food systems*. CRC
17 Press, Taylor & Francis Group, 405 pp. [https://www.crcpress.com/Agroecology-The-Ecology-](https://www.crcpress.com/Agroecology-The-Ecology-of-Sustainable-Food-Systems-Third-Edition/Gliessman/p/book/9781439895610)
18 [of-Sustainable-Food-Systems-Third-Edition/Gliessman/p/book/9781439895610](https://www.crcpress.com/Agroecology-The-Ecology-of-Sustainable-Food-Systems-Third-Edition/Gliessman/p/book/9781439895610) (Accessed
19 March 4, 2019).
- 20 Global Commission on Adaptation, 2019: Global Commission on Adaptation.
21 <https://www.wri.org/our-work/project/global-commission-adaptation> (Accessed April 18, 2019).
- 22 Glotter, M., and J. Elliott, 2016: Simulating US agriculture in a modern Dust Bowl drought. *Nat.*
23 *Plants*, **3**, 16193,
24 doi:10.1038/nplants.2016.193[https://www.nature.com/articles/nplants2016193#supplementary-](https://www.nature.com/articles/nplants2016193#supplementary-information)
25 [information](https://www.nature.com/articles/nplants2016193#supplementary-information). <http://dx.doi.org/10.1038/nplants.2016.193>.
- 26 Godfray, H. C. J., 2015: The debate over sustainable intensification. *Food Secur.*, **7**, 199–208,
27 doi:10.1007/s12571-015-0424-2. <http://link.springer.com/10.1007/s12571-015-0424-2>.
- 28 —, and T. Garnett, 2014: Food security and sustainable intensification. *Philos. Trans. R. Soc. B*
29 *Biol. Sci.*, **369**, 20120273–20120273, doi:10.1098/rstb.2012.0273.
30 <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2012.0273> (Accessed September 18,
31 2018).
- 32 Godfray, H. C. J., and Coauthors, 2010: Food security: the challenge of feeding 9 billion people.
33 *Science*, **327**, 812–818, doi:10.1126/science.1185383.
- 34 Godfray, H. C. J., and Coauthors, 2018: Meat consumption, health, and the environment. *Science (80-*
35 *)*, **361**, eaam5324, doi:10.1126/science.aam5324.
36 <http://www.ncbi.nlm.nih.gov/pubmed/30026199> (Accessed November 4, 2018).
- 37 Goh, A. H. X., 2012: *A Literature Review of the Gender-Differentiated Impacts of Climate Change on*
38 *Women's and Men's Assets and Well-Being in Developing Countries*. Washington, D.C., 43 pp.
39 <http://www.worldagroforestry.org/sites/default/files/4.pdf> (Accessed May 30, 2018).
- 40 Goh, K. M., 2011: Greater Mitigation of Climate Change by Organic than Conventional Agriculture:
41 A Review. *Biol. Agric. Hortic.*, **27**, 205–229, doi:10.1080/01448765.2011.9756648.
42 <http://www.tandfonline.com/doi/abs/10.1080/01448765.2011.9756648> (Accessed March 4,
43 2019).
- 44 Goldstein, B., M. Birkved, J. Fernández, and M. Hauschild, 2017: Surveying the Environmental
45 Footprint of Urban Food Consumption. *J. Ind. Ecol.*, **21**, 151–165, doi:10.1111/jiec.12384.
- 46 González-Varo, J. P., and Coauthors, 2013: Combined effects of global change pressures on animal-
47 mediated pollination. *Trends Ecol. Evol.*, **28**, 524–530, doi:10.1016/j.tree.2013.05.008.

- 1 <http://www.ncbi.nlm.nih.gov/pubmed/23746938> (Accessed October 29, 2018).
- 2 Goulson, D., E. Nicholls, C. Botias, and E. L. Rotheray, 2015: Bee declines driven by combined stress
3 from parasites, pesticides, and lack of flowers. *Science* (80-.), **347**, 1255957–1255957,
4 doi:10.1126/science.1255957. <http://www.ncbi.nlm.nih.gov/pubmed/25721506> (Accessed
5 October 29, 2018).
- 6 Govaerts, B., N. Verhulst, A. Castellanos-Navarrete, K. D. Sayre, J. Dixon, and L. Dendooven, 2009:
7 Conservation Agriculture and Soil Carbon Sequestration: Between Myth and Farmer Reality.
8 *CRC. Crit. Rev. Plant Sci.*, **28**, 97–122, doi:10.1080/07352680902776358.
- 9 Graesser, J., T. M. Aide, H. R. Grau, and N. Ramankutty, 2015: Cropland/pastureland dynamics and
10 the slowdown of deforestation in Latin America. *Environ. Res. Lett.*, **10**, 34017,
11 doi:10.1088/1748-9326/10/3/034017. [http://stacks.iop.org/1748-
12 9326/10/i=3/a=034017?key=crossref.962de1d0c78a7b5affbabdd2f20c7f19](http://stacks.iop.org/1748-9326/10/i=3/a=034017?key=crossref.962de1d0c78a7b5affbabdd2f20c7f19).
- 13 Graham, R. D., and Coauthors, 2007: Nutritious Subsistence Food Systems. *Adv. Agron.*, **92**, 1–74,
14 doi:10.1016/S0065-2113(04)92001-9.
15 <https://www.sciencedirect.com/science/article/pii/S0065211304920019> (Accessed April 5,
16 2019).
- 17 Granderson, A. A., 2014: Making sense of climate change risks and responses at the community level:
18 A cultural-political lens. *Clim. Risk Manag.*, **3**, 55–64, doi:10.1016/J.CRM.2014.05.003.
19 <https://www.sciencedirect.com/science/article/pii/S2212096314000217?via%3Dihub> (Accessed
20 October 23, 2018).
- 21 Gray, C., and V. Mueller, 2012: Drought and Population Mobility in Rural Ethiopia. *World Dev.*, **40**,
22 134–145, doi:10.1016/j.worlddev.2011.05.023.
23 <http://dx.doi.org/10.1016/j.worlddev.2011.05.023>.
- 24 Gray, C. L., 2011: Soil quality and human migration in Kenya and Uganda. *Glob. Environ. Chang.*,
25 **21**, 421–430, doi:10.1016/j.gloenvcha.2011.02.004.
26 <http://dx.doi.org/10.1016/j.gloenvcha.2011.02.004>.
- 27 Gray, C. L., and R. Bilborrow, 2013: Environmental Influences on Human Migration in Rural
28 Ecuador. *Demography*, **50**, 1217–1241, doi:10.1007/s13524-012-0192-y.Environmental.
- 29 von Grebmer, K., and Coauthors, 2014: *2014 Global Hunger Index The Challenge of Hidden Hunger*.
30 Welthungerhilfe, International Food Policy Research Institute, and Concern Worldwide,
31 Washington, D.C., and Dublin, 56 pp.
32 <http://ebrary.ifpri.org/cdm/ref/collection/p15738coll2/id/128360> (Accessed April 20, 2018).
- 33 Grecchi, R. C., Q. H. J. Gwyn, G. B. Bénié, A. R. Formaggio, and F. C. Fahl, 2014: Land use and
34 land cover changes in the Brazilian Cerrado: A multidisciplinary approach to assess the impacts
35 of agricultural expansion. *Appl. Geogr.*, **55**, 300–312, doi:10.1016/J.APGEOG.2014.09.014.
36 <https://www.sciencedirect.com/science/article/pii/S0143622814002240?via%3Dihub> (Accessed
37 May 18, 2018).
- 38 Green, R., J. Milner, A. D. Dangour, A. Haines, Z. Chalabi, A. Markandya, J. Spadaro, and P.
39 Wilkinson, 2015: The potential to reduce greenhouse gas emissions in the UK through healthy
40 and realistic dietary change. *Clim. Change*, **129**, 253–265.
- 41 —, and Coauthors, 2017: Environmental Impacts of Typical Dietary Patterns in India. *FASEB J.*,
42 **31**, 651.2-651.2. http://www.fasebj.org/cgi/content/short/31/1_Supplement/651.2 (Accessed
43 December 3, 2017).
- 44 Green, R. E., S. J. Cornell, J. P. W. Scharlemann, and A. Balmford, 2005: Farming and the fate of
45 wild nature. *Science*, **307**, 550–555, doi:10.1126/science.1106049.
- 46 Greenberg, S. E., R. Bauer, R. Will, R. Locke II, M. Carney, H. Leetaru, and J. Medler, 2017:
47 Geologic Carbon Storage at a One Million Tonne Demonstration Project: Lessons Learned from
48 the Illinois Basin--Decatur Project. *Energy Procedia*, **114**, 5529–5539.

- 1 Griscom, B. W., and Coauthors, 2016: Natural Climate Solutions. *Proc. Natl. Acad. Sci.*, 11–12,
2 doi:10.1073/pnas.1710465114.
- 3 Griscom, B. W., and Coauthors, 2017: Natural climate solutions. *Proc. Natl. Acad. Sci. U. S. A.*, **114**,
4 11645–11650, doi:10.1073/pnas.1710465114. <http://www.ncbi.nlm.nih.gov/pubmed/29078344>
5 (Accessed May 27, 2018).
- 6 Grosso, G., F. Bella, J. Godos, S. Sciacca, D. Del Rio, S. Ray, F. Galvano, and E. L. Giovannucci,
7 2017: Possible role of diet in cancer: systematic review and multiple meta-analyses of dietary
8 patterns, lifestyle factors, and cancer risk. *Nutr. Rev.*, **75**, 405–419, doi:10.1093/nutrit/nux012.
9 [http://academic.oup.com/nutritionreviews/article/75/6/405/3861189/Possible-role-of-diet-in-](http://academic.oup.com/nutritionreviews/article/75/6/405/3861189/Possible-role-of-diet-in-cancer-systematic-review)
10 [cancer-systematic-review](http://academic.oup.com/nutritionreviews/article/75/6/405/3861189/Possible-role-of-diet-in-cancer-systematic-review) (Accessed April 4, 2019).
- 11 Del Grosso, S. J., W. J. Parton, O. Wendroth, R. J. Lascano, and L. Ma, 2019: History of Ecosystem
12 Model Development at Colorado State University and Current Efforts to Address Contemporary
13 Ecological Issues. *Bridging Among Disciplines by Synthesizing Soil and Plant Processes*, Vol.
14 advances in, American Society of Agronomy, Crop Science Society of America, and Soil
15 Science Society of America, Inc.
- 16 Grubler, A., and Coauthors, 2018: A low energy demand scenario for meeting the 1.5 °C target and
17 sustainable development goals without negative emission technologies. *Nat. Energy*, **3**, 515–
18 527, doi:10.1038/s41560-018-0172-6. <http://dx.doi.org/10.1038/s41560-018-0172-6>.
- 19 Guha, I., and C. Roy, 2016: Climate Change, Migration and Food Security: Evidence from Indian
20 Sundarbans. *Int. J. Theor. Appl. Sci.*, **8**, 45–49.
- 21 Gunarathna, M. H. J. P., K. Sakai, T. Nakandakari, M. Kazuro, T. Onodera, H. Kaneshiro, H. Uehara,
22 and K. Wakasugi, 2017: Optimized subsurface irrigation system (OPSIS): Beyond traditional
23 subsurface irrigation. *Water (Switzerland)*, **9**, doi:10.3390/w9080599.
- 24 Gupta, R., E. Somanathan, and S. Dey, 2017: Global warming and local air pollution have reduced
25 wheat yields in India. *Clim. Change*, **140**, 593–604, doi:10.1007/s10584-016-1878-8.
26 <http://link.springer.com/10.1007/s10584-016-1878-8> (Accessed May 17, 2018).
- 27 Gutiérrez-Peña, R., Y. Mena, I. Batalla, and J. M. Mancilla-Leytón, 2019: Carbon footprint of dairy
28 goat production systems: A comparison of three contrasting grazing levels in the Sierra de
29 Grazalema Natural Park (Southern Spain). *J. Environ. Manage.*, **232**, 993–998,
30 doi:10.1016/J.JENVMAN.2018.12.005.
31 <https://www.sciencedirect.com/science/article/pii/S0301479718314142> (Accessed April 16,
32 2019).
- 33 Gutjahr, C., and R. Gerhards, 2010: Decision rules for site-specific weed management. *Precision
34 Crop Protection - The Challenge and Use of Heterogeneity*, E.-C. Oerke, R. Gerhards, G. Menz,
35 and R.A. Sikora, Eds., Springer, Dordrecht, 223–240.
- 36 Haberl, H., T. Beringer, S. C. Bhattacharya, K.-H. Erb, and M. Hoogwijk, 2010: The global technical
37 potential of bio-energy in 2050 considering sustainability constraints. *Curr. Opin. Environ.
38 Sustain.*, **2**, 394–403, doi:10.1016/J.COSUST.2010.10.007.
39 <https://www.sciencedirect.com/science/article/pii/S1877343510001132> (Accessed April 17,
40 2019).
- 41 Hadarits, M., J. Pittman, D. Corkal, H. Hill, K. Bruce, and A. Howard, 2017: The interplay between
42 incremental, transitional, and transformational adaptation: a case study of Canadian agriculture.
43 *Reg. Environ. Chang.*, **17**, 1515–1525, doi:10.1007/s10113-017-1111-y.
- 44 Hadjikakou, M., 2017: Trimming the excess: environmental impacts of discretionary food
45 consumption in Australia. *Ecol. Econ.*, **131**, 119–128, doi:10.1016/j.ecolecon.2016.08.006.
46 <http://dx.doi.org/10.1016/j.ecolecon.2016.08.006>.
- 47 Haines, A., M. Amann, N. Borgford-Parnell, S. Leonard, J. Kuypenstierna, and D. Shindell, 2017:
48 Short-lived climate pollutant mitigation and the Sustainable Development Goals. *Nat. Clim.
49 Chang.*, **7**, 863–869, doi:10.1038/s41558-017-0012-x. <http://www.nature.com/articles/s41558->

- 1 017-0012-x (Accessed March 28, 2019).
- 2 Hallström, E., Q. Gee, P. Scarborough, and D. A. Cleveland, 2017: A healthier US diet could reduce
3 greenhouse gas emissions from both the food and health care systems. *Clim. Change*, **142**, 199–
4 212, doi:10.1007/s10584-017-1912-5.
- 5 Hamilton, A. J., K. Burry, H. F. Mok, S. F. Barker, J. R. Grove, and V. G. Williamson, 2014: Give
6 peas a chance? Urban agriculture in developing countries. A review. *Agron. Sustain. Dev.*, **34**,
7 45–73, doi:10.1007/s13593-013-0155-8.
- 8 Hammill, A., R. Matthew, and E. McCarter, 2010: Microfinance and Climate Change Adaptation. *IDS*
9 *Bull.*, **39**, 113–122, doi:10.1111/j.1759-5436.2008.tb00484.x.
- 10 Hanley, N., T. D. Breeze, C. Ellis, and D. Goulson, 2015: Measuring the economic value of
11 pollination services: Principles, evidence and knowledge gaps. *Ecosyst. Serv.*, **14**, 124–132,
12 doi:http://dx.doi.org/10.1016/j.ecoser.2014.09.013.
13 http://www.sciencedirect.com/science/article/pii/S2212041614001156.
- 14 Hannah, L., and Coauthors, 2017: Regional modeling of climate change impacts on smallholder
15 agriculture and ecosystems in Central America. *Clim. Change*, **141**, 29–45, doi:10.1007/s10584-
16 016-1867-y. http://link.springer.com/10.1007/s10584-016-1867-y (Accessed April 4, 2019).
- 17 Hansen, J., M. Sato, and R. Ruedy, 2012: Perception of climate change. *Proc. Natl. Acad. Sci. U. S.*
18 *A.*, **109**, E2415–E2423, doi:10.1073/pnas.1205276109.
- 19 —, K. Fara, K. Milliken, C. Boyce, L. Chang’a, and E. Allis, 2018: *Strengthening climate services*
20 *for the food security sector*. World Meteorological Organization,
21 https://cgspace.cgiar.org/handle/10568/100218 (Accessed April 17, 2019).
- 22 Harper, A. B., and Coauthors, 2018: Land-use emissions play a critical role in land-based mitigation
23 for Paris climate targets. *Nat. Commun.*, **9**, 2938.
- 24 Harvey, C. A., and Coauthors, 2014a: Climate-Smart Landscapes: Opportunities and Challenges for
25 Integrating Adaptation and Mitigation in Tropical Agriculture. *Conserv. Lett.*, **7**, 77–90,
26 doi:10.1111/conl.12066.
- 27 Harvey, C. A., Z. L. Rakotobe, N. S. Rao, R. Dave, H. Razafimahatratra, R. H. Rabarijohn, H.
28 Rajaofara, and J. L. Mackinnon, 2014b: Extreme vulnerability of smallholder farmers to
29 agricultural risks and climate change in Madagascar. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*,
30 **369**, 20130089, doi:10.1098/rstb.2013.0089. http://www.ncbi.nlm.nih.gov/pubmed/24535397
31 (Accessed October 1, 2018).
- 32 Harwatt, H., J. Sabaté, G. Eshel, S. Soret, and W. Ripple, 2017: Substituting beans for beef as a
33 contribution toward US climate change targets. *Clim. Change*, **143**, 1–10, doi:10.1007/s10584-
34 017-1969-1. http://dx.doi.org/10.1007/s10584-017-1969-1.
- 35 Hasegawa, T., S. Fujimori, K. Takahashi, and T. Masui, 2015: Scenarios for the risk of hunger in the
36 twenty-first century using Shared Socioeconomic Pathways. *Environ. Res. Lett.*, **10**, 14010,
37 doi:10.1088/1748-9326/10/1/014010.
- 38 —, and Coauthors, 2018: Risk of increased food insecurity under stringent global climate change
39 mitigation policy. *Nat. Clim. Chang.*, **8**, 699–703, doi:10.1038/s41558-018-0230-x.
40 http://www.nature.com/articles/s41558-018-0230-x (Accessed August 16, 2018).
- 41 Haughey, E., M. Suter, D. Hofer, N. J. Hoekstra, J. C. McElwain, A. Lüscher, and J. A. Finn, 2018:
42 Higher species richness enhances yield stability in intensively managed grasslands with
43 experimental disturbance. *Sci. Rep.*,.
- 44 Havlík, P., D. Leclère, H. Valin, M. Herrero, E. Schmid, J.-F. Soussana, C. Müller, and M.
45 Obersteiner, 2015: Global climate change, food supply and livestock production systems: A
46 bioeconomic analysis. *Climate change and food systems: global assessments and implications*
47 *for food security and trade*, Food Agriculture Organization of the United Nations (FAO), 176–
48 196.

- 1 Hawken, P., 2017: *Drawdown: the most comprehensive plan ever proposed to reverse global*
2 *warming*. Penguin Books, 240 pp.
- 3 He, J., 2015: Chinese public policy on fisheries subsidies: Reconciling trade, environmental and food
4 security stakes. *Mar. Policy*, **56**, 106–116, doi:10.1016/j.marpol.2014.12.021.
5 <http://dx.doi.org/10.1016/j.marpol.2014.12.021>.
- 6 Heard, B. R., and S. A. Miller, 2016: Critical research needed to examine the environmental impacts
7 of expanded refrigeration on the food system. *Environ. Sci. Technol.*, **50**, 12060–12071.
- 8 Hedenus, F., S. Wirsenius, and D. J. a. Johansson, 2014: The importance of reduced meat and dairy
9 consumption for meeting stringent climate change targets. *Clim. Change*, **124**, 79–91,
10 doi:10.1007/s10584-014-1104-5. <http://link.springer.com/10.1007/s10584-014-1104-5>
11 (Accessed August 31, 2014).
- 12 Hedley, C., 2015: The role of precision agriculture for improved nutrient management on farms. *J.*
13 *Sci. Food Agric.*, doi:10.1002/jsfa.6734.
- 14 Hellberg, R. S., and E. Chu, 2016: Effects of climate change on the persistence and dispersal of
15 foodborne bacterial pathogens in the outdoor environment: A review. *Crit. Rev. Microbiol.*, **42**,
16 548–572, doi:10.3109/1040841x.2014.972335.
- 17 Heller, M., 2019: *Waste Not, Want Not: Reducing Food Loss and Waste in North America Through*
18 *Life Cycle-Based Approaches*. One planet. United Nations Environment Programme (United
19 Nations / intergovernmental organizations), Washington, DC, USA, 83 pp.
20 http://www.oneplanetnetwork.org/sites/default/files/food_loss_and_waste_report.pdf (Accessed
21 April 14, 2019).
- 22 Heller, M. C., and G. A. Keoleian, 2015: Greenhouse Gas Emission Estimates of U.S. Dietary
23 Choices and Food Loss. *J. Ind. Ecol.*, **19**, 391–401, doi:10.1111/jiec.12174.
- 24 Henderson, B., A. Falcucci, A. Mottet, L. Early, B. Werner, H. Steinfeld, and P. Gerber, 2017:
25 Marginal costs of abating greenhouse gases in the global ruminant livestock sector. *Mitig. Adapt.*
26 *Strateg. Glob. Chang.*, **22**, 199–224, doi:10.1007/s11027-015-9673-9.
27 <http://link.springer.com/10.1007/s11027-015-9673-9> (Accessed May 31, 2018).
- 28 Henderson, B., A. Golub, D. Pambudi, T. Hertel, C. Godde, M. Herrero, O. Cacho, and P. Gerber,
29 2018: The power and pain of market-based carbon policies: a global application to greenhouse
30 gases from ruminant livestock production. *Mitig. Adapt. Strateg. Glob. Chang.*, **23**, 349–369,
31 doi:10.1007/s11027-017-9737-0. <http://dx.doi.org/10.1007/s11027-017-9737-0>.
- 32 Henderson, B. B., P. J. Gerber, T. E. Hilinski, A. Falcucci, D. S. Ojima, M. Salvatore, and R. T.
33 Conant, 2015: Greenhouse gas mitigation potential of the world's grazing lands: Modeling soil
34 carbon and nitrogen fluxes of mitigation practices. *Agric. Ecosyst. Environ.*, **207**, 91–100.
- 35 Hendrix, C., and S. Glaser, 2007: Trends and triggers: Climate, climate change and civil conflict in
36 Sub-Saharan Africa. *Polit. Geogr.*, **26**, 695–715,
37 doi:<https://doi.org/10.1016/j.polgeo.2007.06.006>.
- 38 Hendrix, C. S., 2018: Searching for climate–conflict links. *Nat. Clim. Chang.*, **8**, 190–191,
39 doi:10.1038/s41558-018-0083-3. <http://www.nature.com/articles/s41558-018-0083-3> (Accessed
40 May 22, 2018).
- 41 Henry, S., B. Schoumaker, C. Beauchemin, S. Population, and N. May, 2004: The Impact of Rainfall
42 on the First Out-Migration : A Multi-level Event-History Analysis in Burkina Faso All use
43 subject to JSTOR Terms and Conditions The Impact of Rainfall on the First Out-Migration : A
44 Multi-level Event-History Analysis in Burkina Faso. *Popul. Environ.*, **25**, 423–460.
- 45 Herforth, A., and S. Ahmed, 2015: The food environment, its effects on dietary consumption, and
46 potential for measurement within agriculture-nutrition interventions. *Food Secur.*, **7**, 505–520,
47 doi:10.1007/s12571-015-0455-8.
- 48 Hernández-Morcillo, M., J. Hoberg, E. Oteros-Rozas, T. Plieninger, E. Gómez-Baggethun, and V.

- 1 Reyes-García, 2014: Traditional Ecological Knowledge in Europe: Status Quo and Insights for
2 the Environmental Policy Agenda. *Environ. Sci. Policy Sustain. Dev.*, **56**, 3–17,
3 doi:10.1080/00139157.2014.861673.
4 <http://www.tandfonline.com/doi/abs/10.1080/00139157.2014.861673> (Accessed November 3,
5 2018).
- 6 Herrero, M., and Coauthors, 2016a: Greenhouse gas mitigation potentials in the livestock sector. *Nat.*
7 *Clim. Chang.*, **6**, 452–461, doi:10.1038/nclimate2925.
8 <http://www.nature.com/doi/abs/10.1038/nclimate2925> (Accessed December 8, 2017).
- 9 Herrero, M., J. Addison, Bedelian, E. Carabine, P. Havlik, B. Henderson, J. Van de Steeg, and P. K.
10 Thornton, 2016b: Climate change and pastoralism: impacts, consequences and adaptation. *Rev.*
11 *Sci. Tech. l'OIE*, **35**, 417–433, doi:10.20506/rst.35.2.2533.
12 <http://www.ncbi.nlm.nih.gov/pubmed/27917982> (Accessed March 15, 2019).
- 13 Herrero, M., D. Mayberry, J. Van De Steeg, D. Phelan, A. Ash, and K. Diyezee, 2016c:
14 *Understanding Livestock Yield Gaps for Poverty Alleviation , Food Security and the*
15 *Environment*. [https://research.csiro.au/livegaps/wp-content/uploads/sites/37/2015/08/Herrero-et-](https://research.csiro.au/livegaps/wp-content/uploads/sites/37/2015/08/Herrero-et-al.-2016.-Understanding-livestock-yield-gaps-for-poverty-alleviation-food-security-and-the-environment.pdf)
16 [al.-2016.-Understanding-livestock-yield-gaps-for-poverty-alleviation-food-security-and-the-](https://research.csiro.au/livegaps/wp-content/uploads/sites/37/2015/08/Herrero-et-al.-2016.-Understanding-livestock-yield-gaps-for-poverty-alleviation-food-security-and-the-environment.pdf)
17 [environment.pdf](https://research.csiro.au/livegaps/wp-content/uploads/sites/37/2015/08/Herrero-et-al.-2016.-Understanding-livestock-yield-gaps-for-poverty-alleviation-food-security-and-the-environment.pdf).
- 18 Herrero, M. T., and Coauthors, 2010: Climate variability and climate change and their impacts on
19 Kenya’s agricultural sector. <https://cgspace.cgiar.org/handle/10568/3840> (Accessed March 16,
20 2019).
- 21 Heyd, T., and Thomas, 2008: Cultural Responses to Natural Changes such as Climate Change. *Espac.*
22 *Popul. sociétés*, 83–88, doi:10.4000/eps.2397. <http://journals.openedition.org/eps/2397>
23 (Accessed October 23, 2018).
- 24 Hiç, C., P. Pradhan, D. Rybski, and J. P. Kropp, 2016: Food Surplus and Its Climate Burdens.
25 *Environ. Sci. Technol.*, **50**, 4269–4277, doi:10.1021/acs.est.5b05088.
- 26 Hill, S. B., 1985: Redesigning the food system for sustainability. *Alternatives*, **12**, 32–36.
- 27 Hirata, R., and Coauthors, 2013: Carbon dioxide exchange at four intensively managed grassland sites
28 across different climate zones of Japan and the influence of manure application on ecosystem
29 carbon and greenhouse gas budgets. *Agric. For. Meteorol.*, **177**, 57–68,
30 doi:10.1016/J.AGRFORMET.2013.04.007.
31 <https://www.sciencedirect.com/science/article/pii/S0168192313000841> (Accessed April 17,
32 2019).
- 33 HLPE, 2012: *Food security and climate change. A report by the High Level Panel of Experts on Food*
34 *Security and Nutrition of the Committee on World Food Security*. Rome, 102 pp.
35 [http://www.fao.org/fileadmin/user_upload/hlpe/hlpe_documents/HLPE_Reports/HLPE-Report-](http://www.fao.org/fileadmin/user_upload/hlpe/hlpe_documents/HLPE_Reports/HLPE-Report-3-Food_security_and_climate_change-June_2012.pdf)
36 [3-Food_security_and_climate_change-June_2012.pdf](http://www.fao.org/fileadmin/user_upload/hlpe/hlpe_documents/HLPE_Reports/HLPE-Report-3-Food_security_and_climate_change-June_2012.pdf) (Accessed March 19, 2018).
- 37 —, 2014: *Food losses and waste in the context of sustainable food systems. A report by The High*
38 *Level Panel of Experts on Food Security and Nutrition of the Committee on World Food Security*.
39 1–6 pp. www.fao.org/cfs/cfs-hlpe.
- 40 —, 2017: *Nutrition and food systems. A report by the High Level Panel of Experts on Food Security*
41 *and Nutrition of the Committee on World Food Security*. Rome, 152 pp. [http://www.fao.org/3/a-](http://www.fao.org/3/a-i7846e.pdf)
42 [i7846e.pdf](http://www.fao.org/3/a-i7846e.pdf) (Accessed April 4, 2018).
- 43 Hoag, H., 2014: Russian summer tops “universal”heatwave index. *Nature*, **16**.
- 44 Hochman, Z., D. L. Gobbett, and H. Horan, 2017: Climate trends account for stalled wheat yields in
45 Australia since 1990. *Glob. Chang. Biol.*, **23**, 2071–2081, doi:10.1111/gcb.13604.
46 <http://doi.wiley.com/10.1111/gcb.13604> (Accessed May 17, 2018).
- 47 Hodgson, J. A., W. E. Kunin, C. D. Thomas, T. G. Benton, and D. Gabriel, 2010: Comparing organic
48 farming and land sparing: Optimizing yield and butterfly populations at a landscape scale. *Ecol.*

- 1 *Lett.*, **13**, 1358–1367, doi:10.1111/j.1461-0248.2010.01528.x.
- 2 Hofer, D., M. Suter, E. Haughey, J. A. Finn, N. J. Hoekstra, N. Buchmann, and A. Lüscher, 2016:
3 Yield of temperate forage grassland species is either largely resistant or resilient to experimental
4 summer drought. *J. Appl. Ecol.*, **53**, 1023–1034, doi:10.1111/1365-2664.12694.
- 5 Holding, J., J. Carr, and K. Stark, 2013: Food Statistics Pocketbook 2012. 86 pp.
6 https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/183302/foodpocketbook-2012edition-09apr2013.pdf.
- 7
- 8 Hollinger, F., 2015: *Agricultural growth in West Africa market and policy drivers*. FAO,
9 <http://www.fao.org/family-farming/detail/en/c/335194/> (Accessed April 14, 2019).
- 10 Homer-Dixon, T., and Coauthors, 2015: Synchronous failure: The emerging causal architecture of
11 global crisis. *Ecol. Soc.*, **20**, doi:10.5751/ES-07681-200306.
12 <http://www.ecologyandsociety.org/vol20/iss3/art6/>.
- 13 Hossain, M. A., J. Canning, S. Ast, P. J. Rutledge, and A. Jamalipour, 2015: Early warning
14 smartphone diagnostics for water security and analysis using real-time pH mapping. *Photonic
15 Sensors*, **5**, 289–297, doi:10.1007/s13320-015-0256-x.
- 16 Hossain, N., and D. Green, 2011: Living on a Spike: How is the 2011 food price crisis affecting poor
17 people? *Oxfam Policy Pract. Agric. Food L.*, **11**, 9–56.
- 18 Howden, S. M., J.-F. Soussana, F. N. Tubiello, N. Chhetri, M. Dunlop, and H. Meinke, 2007:
19 Adapting agriculture to climate change. *Proc. Natl. Acad. Sci.*, **104**, 19691–19696,
20 doi:10.1073/pnas.0701890104. <http://www.pnas.org/content/104/50/19691.abstract>
21 <http://www.pnas.org/content/104/50/19691.full.pdf>.
- 22 Howell, K. R., P. Shrestha, and I. C. Dodd, 2015: Alternate wetting and drying irrigation maintained
23 rice yields despite half the irrigation volume, but is currently unlikely to be adopted by
24 smallholder lowland rice farmers in Nepal. *Food Energy Secur.*, **4**, 144–157,
25 doi:10.1002/fes3.58. <http://doi.wiley.com/10.1002/fes3.58>.
- 26 Hristov, A. N., and Coauthors, 2013a: SPECIAL TOPICS — Mitigation of methane and nitrous oxide
27 emissions from animal operations: III . A review of animal management mitigation options. *J.
28 Anim. Sci.*, **91**, 5095–5113, doi:10.2527/jas2013-6585.
- 29 —, and Coauthors, 2013b: SPECIAL TOPICS — Mitigation of methane and nitrous oxide
30 emissions from animal operations: I . A review of enteric methane mitigation options. *J. Anim.
31 Sci.*, **91**, 5045–5069, doi:10.2527/jas2013-6583.
- 32 Hristov, A. N., and Coauthors, 2015: An inhibitor persistently decreased enteric methane emission
33 from dairy cows with no negative effect on milk production. *Proc. Natl. Acad. Sci.*, **112**, 10663–
34 10668.
- 35 Hu, Z., J. W. Lee, K. Chandran, S. Kim, and S. K. Khanal, 2012: Nitrous Oxide (N₂O) Emission
36 from Aquaculture: A Review. *Environ. Sci. Technol.*, **46**, 6470–6480, doi:10.1021/es300110x.
37 <http://pubs.acs.org/doi/10.1021/es300110x> (Accessed April 23, 2018).
- 38 —, —, —, —, K. Sharma, A. C. Brotto, and S. K. Khanal, 2013: Nitrogen transformations
39 in intensive aquaculture system and its implication to climate change through nitrous oxide
40 emission. *Bioresour. Technol.*, **130**, 314–320, doi:10.1016/J.BIORTECH.2012.12.033.
41 <https://www.sciencedirect.com/science/article/pii/S0960852412018895> (Accessed April 23,
42 2018).
- 43 Huang, J., and F. Ji, 2015: Effects of climate change on phenological trends and seed cotton yields in
44 oasis of arid regions. *Int. J. Biometeorol.*, **59**, 877–888, doi:10.1007/s00484-014-0904-7.
45 <http://link.springer.com/10.1007/s00484-014-0904-7> (Accessed April 12, 2019).
- 46 Huddleston, B., E. Ataman, P. De Salvo, M. Zanetti, M. Bloise, J. Bel, G. Francheschini, and L.
47 Fed’Ostiani, 2003: *Towards a GIS-based analysis of mountain environments and populations*.
48 FAO Rome, 34 pp pp.

- 1 van Huis, A., 2013: Potential of Insects as Food and Feed in Assuring Food Security. *Annu. Rev.*
2 *Entomol.*, **58**, 563–583, doi:10.1146/annurev-ento-120811-153704.
3 <http://www.annualreviews.org/doi/10.1146/annurev-ento-120811-153704> (Accessed April 17,
4 2019).
- 5 van Huis, A., M. Dicke, and J. J. A. van Loon, 2015: Insects to feed the world. *J. Insects as Food*
6 *Feed*, **1**, 3–5, doi:10.3920/JIFF2015.x002.
7 <https://www.wageningenacademic.com/doi/10.3920/JIFF2015.x002> (Accessed April 17, 2019).
- 8 Hummel, D., 2015: Climate change , environment and migration in the Sahel. *Rural 21*, 40–43.
- 9 Huntington, J. L., K. C. Hegewisch, B. Daudert, C. G. Morton, J. T. Abatzoglou, D. J. McEvoy, and
10 T. Erickson, 2017: Climate engine: Cloud computing and visualization of climate and remote
11 sensing data for advanced natural resource monitoring and process understanding. *Bull. Am.*
12 *Meteorol. Soc.*, **98**, 2397–2409, doi:10.1175/BAMS-D-15-00324.1.
- 13 Hussain, A., G. Rasul, B. Mahapatra, and S. Tuladhar, 2016: Household food security in the face of
14 climate change in the Hindu-Kush Himalayan region. *Food Secur.*, **8**, 921–937,
15 doi:10.1007/s12571-016-0607-5. <http://dx.doi.org/10.1007/s12571-016-0607-5>.
- 16 Huynh, P. T. A., and B. P. Resurreccion, 2014: Women’s differentiated vulnerability and adaptations
17 to climate-related agricultural water scarcity in rural Central Vietnam. *Clim. Dev.*, **6**, 226–237,
18 doi:10.1080/17565529.2014.886989.
19 <http://www.tandfonline.com/doi/abs/10.1080/17565529.2014.886989> (Accessed November 3,
20 2018).
- 21 Iannotti, L. L., M. Robles, H. Pachón, and C. Chiarella, 2012: Food Prices and Poverty Negatively
22 Affect Micronutrient Intakes in Guatemala. *J. Nutr.*, **142**, 1568–1576,
23 doi:10.3945/jn.111.157321. <http://www.ncbi.nlm.nih.gov/pubmed/22695968> (Accessed October
24 21, 2018).
- 25 IAP, 2018: *Opportunities for future research and innovation on food and nutrition security and*
26 *agriculture: The InterAcademy Partnership’s global perspective*. 1-94 pp.
- 27 Ibrahim, M., and A. Kruczkiewicz, 2016: Learning from experience: a review of early warning
28 systems- Moving towards Early Action 2016. 30.
- 29 ICPAC, and WFP, 2017: *Atlas of Climate Risk and Food Security in the Greater Horn of Africa*
30 *Region*. 93 pp pp.
- 31 Ifeanyi-obi, C. C., A. O. Togun, and R. Lamboll, 2016: Influence of Climate Change on Cocoyam
32 Production in Aba Agricultural Zone of Abia State, Nigeria. Springer, Cham, 261–273
33 http://link.springer.com/10.1007/978-3-319-25814-0_18 (Accessed November 4, 2018).
- 34 Iizumi, T., and N. Ramankutty, 2016: Changes in yield variability of major crops for 1981–2010
35 explained by climate change. *Environ. Res. Lett.*, **11**, 34003, doi:10.1088/1748-
36 9326/11/3/034003. [http://stacks.iop.org/1748-
37 9326/11/i=3/a=034003?key=crossref.5949feec63894c33a477f83fc562d2ea](http://stacks.iop.org/1748-9326/11/i=3/a=034003?key=crossref.5949feec63894c33a477f83fc562d2ea) (Accessed April 4,
38 2019).
- 39 —, J. Furuya, Z. Shen, W. Kim, M. Okada, S. Fujimori, T. Hasegawa, and M. Nishimori, 2017:
40 Responses of crop yield growth to global temperature and socioeconomic changes. *Sci. Rep.*, **7**,
41 7800, doi:10.1038/s41598-017-08214-4.
42 <http://www.ncbi.nlm.nih.gov/pmc/articles/PMC5552729/>.
- 43 —, H. Shiogama, Y. Imada, N. Hanasaki, H. Takikawa, and M. Nishimori, 2018: Crop production
44 losses associated with anthropogenic climate change for 1981-2010 compared with preindustrial
45 levels. *Int. J. Climatol.*, doi:10.1002/joc.5818. <http://doi.wiley.com/10.1002/joc.5818> (Accessed
46 October 18, 2018).
- 47 Imamura, F., R. Micha, S. Khatibzadeh, S. Fahimi, P. Shi, J. Powles, and D. Mozaffarian, 2015:
48 Dietary quality among men and women in 187 countries in 1990 and 2010: A systematic

- 1 assessment. *Lancet Glob. Heal.*, **3**, e132–e142, doi:10.1016/S2214-109X(14)70381-X.
- 2 Indoria, A. K., C. Srinivasa Rao, K. L. Sharma, and K. Sammi Reddy, 2017: Conservation agriculture
3 - a panacea to improve soil physical health. *Curr. Sci.*, **112**, 52–61,
4 doi:10.18520/cs/v112/i01/52-61.
- 5 Innes, P. J., D. K. Y. Tan, F. Van Ogtrop, and J. S. Amthor, 2015: Effects of high-temperature
6 episodes on wheat yields in New South Wales, Australia. *Agric. For. Meteorol.*, **208**, 95–107,
7 doi:10.1016/J.AGRFORMET.2015.03.018.
8 <https://www.sciencedirect.com/science/article/pii/S0168192315000969> (Accessed April 4,
9 2019).
- 10 INPE, 2015: *Nota tecnica: INPE apresenta taxa de desmatamento consolidada do PRODES 2015.*
11 Brazil, 3 pp pp. http://www.obt.inpe.br/prodes/Prodes_Taxa2015_consolidada.pdf (Accessed
12 September 19, 2018).
- 13 —, 2018: Projeto de Monitoramento do Desmatamento na Amazônia Legal - PRODES (INPE).
14 *Natl. Inst. Sp. Res.*, <http://www.obt.inpe.br/prodes/index.php> (Accessed September 18, 2018).
- 15 International Council for Science (ICSU), 2017: *A Guide to SDG Interactions: from Science to*
16 *Implementation*. Paris, 239 pp pp.
- 17 IPBES, 2016: *Summary for policymakers of the assessment report of the Intergovernmental Science-*
18 *Policy Platform on Biodiversity and Ecosystem Services on pollinators, pollination and food*
19 *production*. Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services
20 (IPBES), Bonn, Germany, 1-30 pp. <http://centaur.reading.ac.uk/69028/> (Accessed November 4,
21 2018).
- 22 IPCC, 2007: *Climate Change 2007: Impacts, Adaptation and Vulnerability. Contribution of Working*
23 *Group II to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change.*
24 M.L. Parry, O.F. Canziani, J.P. Palutikof, P.J. Van der Linden, and C.E. Hanson, Eds.
25 Cambridge University Press, Cambridge, UK, 976 pp.
26 https://www.ipcc.ch/site/assets/uploads/2018/03/ar4_wg2_full_report.pdf.
- 27 —, 2012: *Managing the Risks of Extreme Events and Disasters to Advance Climate Change*
28 *Adaptation. A Special Report of Working Groups I and II of the Intergovernmental Panel on*
29 *Climate Change*. [Field, C.B., V. Barros, T.F. Stocker, D. Qin, D.J. Dokken, K.L. Ebi, M.D.
30 Mastrandrea, K.J. Mach, G.-K. Plattner, S.K. Allen, M. Tignor, and P.M. Midgley (eds.)].
31 Cambridge, UK, and New York, NY, USA, 582 pp. [https://www.ipcc.ch/report/managing-the-](https://www.ipcc.ch/report/managing-the-risks-of-extreme-events-and-disasters-to-advance-climate-change-adaptation/)
32 [risks-of-extreme-events-and-disasters-to-advance-climate-change-adaptation/](https://www.ipcc.ch/report/managing-the-risks-of-extreme-events-and-disasters-to-advance-climate-change-adaptation/) (Accessed March
33 7, 2019).
- 34 —, 2014a: *Climate Change 2014: Mitigation of Climate Change: Working Group III Contribution*
35 *to the IPCC Fifth Assessment Report*. Cambridge University Press, Cambridge, 1435 pp.
36 [https://www.cambridge.org/core/books/climate-change-2014-mitigation-of-climate-](https://www.cambridge.org/core/books/climate-change-2014-mitigation-of-climate-change/81F2F8D8D234727D153EC10D428A2E6D)
37 [change/81F2F8D8D234727D153EC10D428A2E6D](https://www.cambridge.org/core/books/climate-change-2014-mitigation-of-climate-change/81F2F8D8D234727D153EC10D428A2E6D).
- 38 —, 2014b: *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to*
39 *the Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Core Writing*
40 *Team, R.K. Pachauri and L.A. Meyer (eds.)]*. IPCC, Geneva,.
- 41 —, 2014c: *Climate Change 2014: Impacts, Adaptation, and Vulnerability: Working Group II*
42 *Contribution to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change.*
43 Cambridge, <https://www.ipcc.ch/report/ar5/wg2/>.
- 44 —, 2018a: *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of*
45 *1.5 °C above pre-industrial levels and related global greenhouse gas emission pathways, in the*
46 *context of strengthening the global response to the threat of climate change.*
47 <http://www.ipcc.ch/report/sr15/>.
- 48 —, 2018b: Summary for Policymakers. *Global Warming of 1.5 °C an IPCC special report on the*
49 *impacts of global warming of 1.5 °C above pre-industrial levels and related global greenhouse*

- 1 *gas emission pathways, in the context of strengthening the global response to the threat of*
2 *climate change*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge, United
3 Kingdom and New York, NY, USA. http://report.ipcc.ch/sr15/pdf/sr15_spm_final.pdf (Accessed
4 October 29, 2018).
- 5 IRD, 2017: *Rural societies in the face of climatic and environmental changes in West Africa*. S.M.A.
6 (ed. . Sultan Benjamin (ed.), Lalou Richard (ed.), Amadou Sanni M. (ed.), Oumarou A. (ed.),
7 Ed. Marseille, 432 pp. <http://www.editions.ird.fr/produit/462/9782709924269> (Accessed April
8 14, 2019).
- 9 Isbell, F., and Coauthors, 2017: Benefits of increasing plant diversity in sustainable agroecosystems.
10 *J. Ecol.*, doi:10.1111/1365-2745.12789.
- 11 Ishigooka, Y., S. Fukui, T. Hasegawa, T. Kuwagata, M. Nishimori, and M. Kondo, 2017: Large-scale
12 evaluation of the effects of adaptation to climate change by shifting transplanting date on rice
13 production and quality in Japan. *J. Agric. Meteorol.*, **73**, 156–173, doi:10.2480/agrmet.D-16-
14 00024.
- 15 Jaberalansar, Z., M. Tarkesh, M. Bassiri, and S. Pourmanafi, 2017: Modelling the impact of climate
16 change on rangeland forage production using a generalized regression neural network: a case
17 study in Isfahan Province, Central Iran. *J. Arid Land*, **9**, 489–503, doi:10.1007/s40333-017-
18 0058-7. <http://link.springer.com/10.1007/s40333-017-0058-7> (Accessed April 25, 2019).
- 19 Jägermeyr, J., D. Gerten, S. Schaphoff, J. Heinke, W. Lucht, and J. Rockström, 2016: Integrated crop
20 water management might sustainably halve the global food gap. *Environ. Res. Lett.*, **11**, 25002,
21 doi:10.1088/1748-9326/11/2/025002. [http://iopscience.iop.org/article/10.1088/1748-
22 9326/11/2/025002%5Cnhttp://stacks.iop.org/1748-
23 9326/11/i=2/a=025002?key=crossref.bc805e58c0fb73d1402f4ae22c75639a](http://iopscience.iop.org/article/10.1088/1748-9326/11/2/025002%5Cnhttp://stacks.iop.org/1748-9326/11/2/025002?key=crossref.bc805e58c0fb73d1402f4ae22c75639a).
- 24 Jägermeyr, J., A. Pastor, H. Biemans, and D. Gerten, 2017: Reconciling irrigated food production
25 with environmental flows for Sustainable Development Goals implementation. *Nat. Commun.*,
26 **8**, 15900, doi:10.1038/ncomms15900. <http://www.nature.com/doifinder/10.1038/ncomms15900>.
- 27 Jara-Samaniego, J., and Coauthors, 2017: Development of organic fertilizers from food market waste
28 and urban gardening by composting in Ecuador. *PLoS One*, **12**, e0181621,
29 doi:10.1371/journal.pone.0181621. <https://dx.plos.org/10.1371/journal.pone.0181621> (Accessed
30 April 14, 2019).
- 31 Jat, M. L., Yadvinder-Singh, G. Gill, H. S. Sidhu, J. P. Aryal, C. Stirling, and B. Gerard, 2015: Laser-
32 Assisted Precision Land Leveling Impacts in Irrigated Intensive Production Systems of South
33 Asia. *Advances in Soil Science*, 323–352.
- 34 ———, and Coauthors, 2016: Climate change and agriculture: Adaptation strategies and mitigation
35 opportunities for food security in South Asia and Latin America. *Advances in Agronomy*, Vol.
36 137 of, 127–235.
- 37 Jat, R. K., T. B. Sapkota, R. G. Singh, M. L. Jat, M. Kumar, and R. K. Gupta, 2014: Seven years of
38 conservation agriculture in a rice–wheat rotation of Eastern Gangetic Plains of South Asia: Yield
39 trends and economic profitability. *F. Crop. Res.*, doi:<http://dx.doi.org/10.1016/j.fcr.2014.04.015>.
- 40 Jensen, H. G., L. B. Jacobsen, S. M. Pedersen, and E. Tavella, 2012: Socioeconomic impact of
41 widespread adoption of precision farming and controlled traffic systems in Denmark. *Precis.*
42 *Agric.*, doi:10.1007/s11119-012-9276-3.
- 43 Jessoe, K., D. T. Manning, and J. E. Taylor, 2018: Climate Change and Labour Allocation in Rural
44 Mexico: Evidence from Annual Fluctuations in Weather. *Econ. J.*, **128**, 230–261,
45 doi:10.1111/eoj.12448. <https://academic.oup.com/ej/article/128/608/230-261/5068979>
46 (Accessed March 20, 2019).
- 47 Jha, B., and A. Tripathi, 2017: How Susceptible Is India’s Food Basket to Climate Change? *Soc.*
48 *Change*, **47**, 11–27.

- 1 Johansson, D. J. A., and C. Azar, 2007: A scenario based analysis of land competition between food
2 and bioenergy production in the US. *Clim. Change*, **82**, 267–291, doi:10.1007/s10584-006-9208-
3 1.
- 4 Jones, A., and B. Hiller, 2017: Exploring the Dynamics of Responses to Food Production Shocks.
5 *Sustainability*, **9**, 960. <http://www.mdpi.com/2071-1050/9/6/960>.
- 6 Jones, A. D., 2017: On-Farm Crop Species Richness Is Associated with Household Diet Diversity and
7 Quality in Subsistence- and Market-Oriented Farming Households in Malawi. *J. Nutr.*, **147**, 86–
8 96, doi:10.3945/jn.116.235879. <https://academic.oup.com/jn/article/147/1/86-96/4584667>
9 (Accessed October 29, 2018).
- 10 —, F. M. Ngunjiri, G. Pelto, and S. L. Young, 2013: What Are We Assessing When We Measure
11 Food Security? A Compendium and Review of Current Metrics 1, 2. *Am. Soc. Nutr. Adv. Nutr.*,
12 481–505, doi:10.3945/an.113.004119.disciplines.
- 13 Jones, A. D., A. Shrinivas, and R. Bezner-Kerr, 2014: Farm production diversity is associated with
14 greater household dietary diversity in Malawi: Findings from nationally representative data.
15 *Food Policy*, **46**, 1–12, doi:10.1016/J.FOODPOL.2014.02.001.
16 <https://www.sciencedirect.com/science/article/abs/pii/S0306919214000256> (Accessed October
17 27, 2018).
- 18 Jost, C., and Coauthors, 2016: Understanding gender dimensions of agriculture and climate change in
19 smallholder farming communities. *Clim. Dev.*, **8**, 133–144,
20 doi:10.1080/17565529.2015.1050978.
- 21 Jung, S., L. V. Rasmussen, C. Watkins, P. Newton, and A. Agrawal, 2017: Brazil’s National
22 Environmental Registry of Rural Properties: Implications for Livelihoods. *Ecol. Econ.*, **136**, 53–
23 61, doi:10.1016/j.ecolecon.2017.02.004.
24 <https://linkinghub.elsevier.com/retrieve/pii/S0921800916308758>.
- 25 Juroszek, P., and A. von Tiedemann, 2013: Climate change and potential future risks through wheat
26 diseases: a review. *Eur. J. Plant Pathol.*, **136**, 21–33, doi:10.1007/s10658-012-0144-9.
- 27 Kabir, M., S. Rafiq, S. M. K. Salema, and H. Scheyvens, 2016: Approaches of MFIs to Disasters and
28 Climate Change Adaptation in Bangladesh Working Paper No . 49 Approaches of MFIs to
29 Disasters and Climate Change Adaptation in Bangladesh. *Inst. Incl. Financ. Dev. (InM), Work.*
30 *Pap., Working Pa.*
- 31 Kadim, I. T., O. Mahgoub, S. Baqir, B. Faye, and R. Purchas, 2015: Cultured meat from muscle stem
32 cells: A review of challenges and prospects. *J. Integr. Agric.*, **14**, 222–233, doi:10.1016/S2095-
33 3119(14)60881-9. <https://www.sciencedirect.com/science/article/pii/S2095311914608819>
34 (Accessed April 23, 2019).
- 35 Kahane, R., and Coauthors, 2013: Agrobiodiversity for food security, health and income. *Agron.*
36 *Sustain. Dev.*, **33**, 671–693, doi:10.1007/s13593-013-0147-8.
37 <http://link.springer.com/10.1007/s13593-013-0147-8> (Accessed March 24, 2019).
- 38 Kaijser, A., and A. Kronsell, 2014: Climate change through the lens of intersectionality. *Env. Polit.*,
39 **23**, 417–433, doi:10.1080/09644016.2013.835203.
40 <http://www.tandfonline.com/doi/abs/10.1080/09644016.2013.835203> (Accessed October 28,
41 2018).
- 42 Kantachote, D., T. Nunkaew, T. Kantha, and S. Chaiprapat, 2016: Biofertilizers from
43 *Rhodopseudomonas palustris* strains to enhance rice yields and reduce methane emissions. *Appl.*
44 *Soil Ecol.*, **100**, 154–161, doi:10.1016/j.apsoil.2015.12.015.
45 <http://dx.doi.org/10.1016/j.apsoil.2015.12.015>.
- 46 Kanter, R., and B. Caballero, 2012: Global gender disparities in obesity: a review. *Adv. Nutr.*, **3**, 491–
47 498, doi:10.3945/an.112.002063. <http://www.ncbi.nlm.nih.gov/pubmed/22797984> (Accessed
48 December 8, 2017).

- 1 Karimi, V., E. Karami, and M. Keshavarz, 2018: Climate change and agriculture: Impacts and
2 adaptive responses in Iran. *J. Integr. Agric.*, **17**, 1–15, doi:10.1016/S2095-3119(17)61794-5.
3 <https://www.sciencedirect.com/science/article/pii/S2095311917617945> (Accessed April 24,
4 2019).
- 5 Kawanishi, M., B. L. Preston, and N. A. Ridwan, 2016: Evaluation of National Adaptation Planning:
6 A Case Study in Indonesia. *Climate Change Policies and Challenges in Indonesia*.
- 7 Kehlbacher, A., R. Tiffin, A. Briggs, M. Berners-Lee, and P. Scarborough, 2016: The distributional
8 and nutritional impacts and mitigation potential of emission-based food taxes in the UK. *Clim.*
9 *Change*, **137**, 121–141, doi:10.1007/s10584-016-1673-6.
10 <http://link.springer.com/10.1007/s10584-016-1673-6> (Accessed April 5, 2019).
- 11 Keles, D., J. Choumert-Nkolo, P. Combes Motel, and E. Nazindigouba Kéré, 2018: Does the
12 expansion of biofuels encroach on the forest? *J. For. Econ.*, **33**, 75–82,
13 doi:10.1016/J.JFE.2018.11.001.
- 14 Kelley, C., S. Mohtadi, M. Cane, R. Seager, and Y. Kushnir, 2017: Commentary on the Syria case:
15 Climate as a contributing factor. *Polit. Geogr.*, **60**, 245–247, doi:10.1016/j.polgeo.2017.06.013.
16 <http://dx.doi.org/10.1016/j.polgeo.2017.06.013>.
- 17 Kelley, C. P., S. Mohtadi, M. A. Cane, R. Seager, and Y. Kushnir, 2015: Climate change in the Fertile
18 Crescent and implications of the recent Syrian drought. *Proc. Natl. Acad. Sci.*, **112**, 3241–3246,
19 doi:10.1073/pnas.1421533112. <http://www.pnas.org/lookup/doi/10.1073/pnas.1421533112>.
- 20 Kemper, J., 2015: Biomass and carbon dioxide capture and storage: A review. *Int. J. Greenh. Gas*
21 *Control*, **40**, 401–430.
- 22 Kench, P. S., M. R. Ford, and S. D. Owen, 2018: Patterns of island change and persistence offer
23 alternate adaptation pathways for atoll nations. *Nat. Commun.*, **9**, 605, doi:10.1038/s41467-018-
24 02954-1. <http://www.nature.com/articles/s41467-018-02954-1> (Accessed May 8, 2018).
- 25 Kerr, R. B., E. Chilanga, H. Nyantakyi-Frimpong, I. Luginaah, and E. Lupafya, 2016: Integrated
26 agriculture programs to address malnutrition in northern Malawi. *BMC Public Health*, **16**, 1197,
27 doi:10.1186/s12889-016-3840-0. <http://www.ncbi.nlm.nih.gov/pubmed/27894303> (Accessed
28 October 28, 2018).
- 29 Kershen, D. L., 2013: The Contested Vision for Agriculture’s Future: Sustainable Intensive
30 Agriculture and Agroecology Agriculture. *Creighton Law Rev.*, **46**, 591–618.
31 [https://heinonline.org/HOL/Page?handle=hein.journals/creigh46&id=625&collection=journals&
32 index=](https://heinonline.org/HOL/Page?handle=hein.journals/creigh46&id=625&collection=journals&index=) (Accessed April 13, 2019).
- 33 Kerton, S., and A. J. Sinclair, 2010: Buying local organic food: a pathway to transformative learning.
34 *Agric. Human Values*, **27**, 401–413, doi:10.1007/s10460-009-9233-6.
35 <http://link.springer.com/10.1007/s10460-009-9233-6> (Accessed March 25, 2019).
- 36 Ketiemi, P., P. M. Makeni, E. K. Maranga, and P. A. Omondi, 2017: Integration of climate change
37 information into drylands crop production practices for enhanced food security: A case study of
38 Lower Tana Basin in Kenya. *African J. Agric. Res.*, **12**, 1763–1771.
- 39 Khanal, R., A. Timilsina, P. Pokhrel, and R. K. P. Yadav, 2015: Documenting Abundance and Use of
40 Underutilized Plant Species in the Mid Hill Region of Nepal. *Ecoprint An Int. J. Ecol.*, **21**, 63–
41 71, doi:10.3126/eco.v21i0.11906.
- 42 Kiefer, I., T. Rathmann, and M. Kunze, 2005: Eating and dieting differences in men and women. *J.*
43 *Men’s Heal. Gen.*, **2**, 194–201, doi:10.1016/j.jmhg.2005.04.010.
44 <http://online.liebertpub.com/doi/pdfplus/10.1016/j.jmhg.2005.04.010> (Accessed April 5, 2019).
- 45 Kihara, J., and Coauthors, 2015: Perspectives on climate effects on agriculture: The international
46 efforts of AgMIP in Sub-Saharan Africa. *Handbook of Climate Change and agroecosystems:
47 The Agricultural Model Intercomparison and Improvement Project Integrated Crop and
48 Economic Assessments, Part 2*, World Scientific, 3–23.

- 1 Kim, K. H., and J. Cho, 2016: Predicting potential epidemics of rice diseases in Korea using multi-
2 model ensembles for assessment of climate change impacts with uncertainty information. *Clim.*
3 *Change*, **134**, 327–339, doi:10.1007/s10584-015-1503-2.
- 4 ———, ———, Y. H. Lee, and W. S. Lee, 2015: Predicting potential epidemics of rice leaf blast and
5 sheath blight in South Korea under the RCP 4.5 and RCP 8.5 climate change scenarios using a
6 rice disease epidemiology model, EPIRICE. *Agric. For. Meteorol.*, **203**, 191–207,
7 doi:10.1016/j.agrformet.2015.01.011.
- 8 Kim, T. J., and O. von dem Knesebeck, 2018: Income and obesity: what is the direction of the
9 relationship? A systematic review and meta-analysis. *BMJ Open*, **8**.
10 <http://bmjopen.bmj.com/content/8/1/e019862.abstract>.
- 11 Kirezieva, K., L. Jacxsens, M. van Boekel, and P. A. Luning, 2015: Towards strategies to adapt to
12 pressures on safety of fresh produce due to climate change. *Food Res. Int.*, **68**, 94–107,
13 doi:10.1016/j.foodres.2014.05.077.
- 14 Klein, A. M., B. E. Vaissiere, J. H. Cane, I. Steffan-Dewenter, S. A. Cunningham, C. Kremen, and T.
15 Tscharntke, 2007: Importance of pollinators in changing landscapes for world crops. *Proc. R.*
16 *Soc. B-Biological Sci.*, **274**, 303–313, doi:10.1098/rspb.2006.3721.
- 17 Klenk, N., A. Fiume, K. Meehan, and C. Gibbes, 2017: Local knowledge in climate adaptation
18 research: moving knowledge frameworks from extraction to co-production. *Wiley Interdiscip.*
19 *Rev. Clim. Chang.*, **8**, doi:10.1002/wcc.475.
- 20 Köberle, A. C., I. Schmidt Tagomori, J. Portugal-Pereira, and R. Schaeffer, 2016: *Policy Case Study:*
21 *Evaluation of the Plano ABC in Brazil*. Brazil, 105-133 pp.
- 22 Köberle, A. C., J. Portugal-Pereira, B. Cunha, R. Garaffa, A. F. Lucena, A. Szklo, and R. Schaeffer,
23 2019: Brazilian ethanol expansion subject to limitations. *Nat. Clim. Chang.*, 209.
- 24 Kole, C., M. Muthamilarasan, R. Henry, and D. Edwards, 2015: Application of genomics-assisted
25 breeding for generation of climate resilient crops: progress and prospects. *Front. Plant Sci.*, **6**,
26 1–16, doi:10.3389/fpls.2015.00563.
27 <http://journal.frontiersin.org/Article/10.3389/fpls.2015.00563/abstract>.
- 28 Koohafkan, P., and M. A. Altieri, 2011: *Globally important agricultural heritage systems: a legacy*
29 *for the future*. Food and Agriculture Organization of the United Nations Rome, Rome, Italy, 1-
30 49 pp.
- 31 Korbel'ová, L., and S. Kohnová, 2017: Methods for Improvement of the Ecosystem Services of Soil
32 by Sustainable Land Management in the Myjava River Basin. *Slovak J. Civ. Eng.*, **25**, 29–36,
33 doi:10.1515/sjce-2017-0005. [http://www.degruyter.com/view/j/sjce.2017.25.issue-1/sjce-2017-](http://www.degruyter.com/view/j/sjce.2017.25.issue-1/sjce-2017-0005/sjce-2017-0005.xml)
34 [0005/sjce-2017-0005.xml](http://www.degruyter.com/view/j/sjce.2017.25.issue-1/sjce-2017-0005/sjce-2017-0005.xml).
- 35 Kraxner, F., and Coauthors, 2014: BECCS in South Korea—analyzing the negative emissions
36 potential of bioenergy as a mitigation tool. *Renew. Energy*, **61**, 102–108.
- 37 Kremen, C., A. Iles, and C. Bacon, 2012: Diversified Farming Systems: An Agroecological, Systems-
38 based Alternative to Modern Industrial Agriculture. *Ecol. Soc.*, **17**, art44, doi:10.5751/ES-
39 05103-170444. <http://www.ecologyandsociety.org/vol17/iss4/art44/> (Accessed March 4, 2019).
- 40 Kriewald, S., P. Pradhan, L. Costa, R. A. Cantu, and J. Kropp, Hungry cities: how local food self-
41 sufficiency relates to climate change, life styles and demographic development. *Environ. Res.*
42 *Lett.*,.
- 43 Krishnapillai, M., 2018: Enhancing Adaptive Capacity and Climate Change Resilience of Coastal
44 Communities in Yap. *Climate Change Impacts and Adaptation Strategies for Coastal*
45 *Communities*, W.L. Filho, Ed., Springer International Publishing AG, Cham, 87–118
46 http://link.springer.com/10.1007/978-3-319-70703-7_5 (Accessed January 12, 2018).
- 47 Krishnapillai, M., and R. Gavenda, 2014: *From barren land to biodiverse home gardens*. 26-28 pp.

- 1 Krishnapillai, M. V, 2017: Climate-Friendly Adaptation Strategies for the Displaced Atoll Population
2 in Yap. *Climate Change Adaptation in Pacific Countries: Fostering Resilience and Improving*
3 *the Quality of Life*, W. Leal Filho, Ed., Springer International Publishing, Cham, 101–117
4 https://doi.org/10.1007/978-3-319-50094-2_6.
- 5 Kucich, D. A., and M. M. Wicht, 2016: South African indigenous fruits – Underutilized resource for
6 boosting daily antioxidant intake among local indigent populations? *South African J. Clin. Nutr.*,
7 **29**, 150–156, doi:10.1080/16070658.2016.1219470.
8 <https://www.tandfonline.com/doi/full/10.1080/16070658.2016.1219470> (Accessed October 21,
9 2018).
- 10 Kumar, N., and A. R. Quisumbing, 2013: Gendered impacts of the 2007–2008 food price crisis:
11 Evidence using panel data from rural Ethiopia. *Food Policy*, **38**, 11–22,
12 doi:10.1016/J.FOODPOL.2012.10.002.
13 <https://www.sciencedirect.com/science/article/abs/pii/S0306919212001029> (Accessed
14 November 3, 2018).
- 15 —, J. Harris, and R. Rawat, 2015: If They Grow It, Will They Eat and Grow? Evidence from
16 Zambia on Agricultural Diversity and Child Undernutrition. *J. Dev. Stud.*, **51**, 1060–1077,
17 doi:10.1080/00220388.2015.1018901.
18 <http://www.tandfonline.com/doi/full/10.1080/00220388.2015.1018901> (Accessed October 28,
19 2018).
- 20 Kumar, R., V. Mishra, J. Buzan, R. Kumar, D. Shindell, and M. Huber, 2017a: Dominant control of
21 agriculture and irrigation on urban heat island in India. *Sci. Rep.*, **7**, 14054, doi:10.1038/s41598-
22 017-14213-2. <https://doi.org/10.1038/s41598-017-14213-2>.
- 23 Kumar, Y., R. Berwal, A. Pandey, A. Sharma, and V. Sharma, 2017b: Hydroponics meat: An
24 envisaging boon for sustainable meat production through biotechnological approach - A review.
25 *Int. J. Vet. Sci. Anim. Husb.*, **2**, 34–39. [http://www.veterinarypaper.com/archives/2017/2/1/A/1-
26 3-25](http://www.veterinarypaper.com/archives/2017/2/1/A/1-3-25) (Accessed November 4, 2018).
- 27 Läderach, P., J. R. Villegas, C. Navarro-racines, C. Zelaya, A. M. Valle, and A. Jarvis, 2017: Climate
28 change adaptation of coffee production in space and time. *Clim. Change*, **141**, 47–62,
29 doi:10.1007/s10584-016-1788-9.
- 30 Lal, R., 2004: Soil Carbon Sequestration Impacts on Global Climate Change and Food Security.
31 *Science (80-.)*, **304**, 1623–1627.
- 32 Lal, S., and Coauthors, 2017: Loss of crop yields in India due to surface ozone: an estimation based
33 on a network of observations. *Environ. Sci. Pollut. Res.*, **24**, 20972–20981, doi:10.1007/s11356-
34 017-9729-3. <http://link.springer.com/10.1007/s11356-017-9729-3> (Accessed April 25, 2019).
- 35 Lambin, E. F., and Coauthors, 2013: Estimating the world’s potentially available cropland using a
36 bottom-up approach. *Glob. Environ. Chang.*, **23**, 892–901,
37 doi:10.1016/j.gloenvcha.2013.05.005.
- 38 Lambin, E. F., and Coauthors, 2018: The role of supply-chain initiatives in reducing deforestation.
39 *Nat. Clim. Chang.*, **8**, 109–116, doi:10.1038/s41558-017-0061-1.
40 <http://www.nature.com/articles/s41558-017-0061-1>.
- 41 Lamers, P., C. Hamelinck, M. Junginger, and A. Faaij, 2011: International bioenergy trade—A review
42 of past developments in the liquid biofuel market. *Renew. Sustain. Energy Rev.*, **15**, 2655–2676,
43 doi:10.1016/J.RSER.2011.01.022.
- 44 Lamichhane, J. R., and Coauthors, 2015: Robust cropping systems to tackle pests under climate
45 change. A review. *Agron. Sustain. Dev.*, **35**, 443–459, doi:10.1007/s13593-014-0275-9.
- 46 Lampkin, N., and Coauthors, 2015: The Role of Agroecology in Sustainable Intensification.
47 <http://orgprints.org/33067/> (Accessed March 3, 2019).
- 48 Landholm, D. M., P. Pradhan, and J. P. Kropp, 2019: Diverging forest land use dynamics induced by

- 1 armed conflict across the tropics. *Glob. Environ. Chang.*, **56**, 86–94,
2 doi:10.1016/j.gloenvcha.2019.03.006.
- 3 Langer, A., and Coauthors, 2015: Women and Health: the key for sustainable development. *Lancet*
4 (London, England), **386**, 1165–1210, doi:10.1016/S0140-6736(15)60497-4.
5 <http://www.ncbi.nlm.nih.gov/pubmed/26051370> (Accessed March 13, 2019).
- 6 Lara, L., and M. Rostagno, 2013: Impact of Heat Stress on Poultry Production. *Animals*, **3**, 356.
7 <http://www.mdpi.com/2076-2615/3/2/356>.
- 8 Larkin, A., J. Kuriakose, M. Sharmina, and K. Anderson, 2018: What if negative emission
9 technologies fail at scale? Implications of the Paris Agreement for big emitting nations. *Clim.*
10 *Policy*, **18**, 690–714.
- 11 Latham, A. D. M., M. C. Latham, E. Cieraad, D. M. Tompkins, and B. Warburton, 2015: Climate
12 change turns up the heat on vertebrate pest control. *Biol. Invasions*, **17**, 2821–2829,
13 doi:10.1007/s10530-015-0931-2. <http://link.springer.com/10.1007/s10530-015-0931-2>
14 (Accessed March 4, 2019).
- 15 Laue, J. E., and E. Y. Arima, 2016: Spatially explicit models of land abandonment in the Amazon. *J.*
16 *Land Use Sci.*, **11**, 48–75, doi:10.1080/1747423X.2014.993341.
- 17 Lebel, L., P. Lebel, and B. Lebel, 2014: Gender and the management of climate-related risks in
18 Northern Thailand. *Int. Soc. Sci. J.*, **65**, 147–158, doi:10.1111/issj.12090.
19 <http://doi.wiley.com/10.1111/issj.12090> (Accessed October 29, 2018).
- 20 Lehmann, J., J. Gaunt, and M. Rondon, 2006: Bio-char Sequestration in Terrestrial Ecosystems – A
21 Review. *Mitig. Adapt. Strateg. Glob. Chang.*, **11**, 403–427, doi:10.1007/s11027-005-9006-5.
22 <http://link.springer.com/10.1007/s11027-005-9006-5> (Accessed November 5, 2018).
- 23 ———, Y. Kuzyakov, G. Pan, and Y. S. Ok, 2015: Biochars and the plant-soil interface. *Plant Soil*, **395**,
24 1–5, doi:10.1007/s11104-015-2658-3. <http://link.springer.com/10.1007/s11104-015-2658-3>
25 (Accessed November 5, 2018).
- 26 Lenton, T. M., 2010: The potential for land-based biological CO₂ removal to lower future
27 atmospheric CO₂ concentration. *Carbon Manag.*, **1**, 145–160, doi:10.4155/cmt.10.12.
28 <http://www.tandfonline.com/doi/abs/10.4155/cmt.10.12> (Accessed April 25, 2019).
- 29 Lentz, E., C. B. Barrett, and J. Hoddinott, 2005: *Food Aid and Dependency: Implications for*
30 *Emergency Food Security Assessments*. 1-50 pp.
31 https://papers.ssrn.com/sol3/papers.cfm?abstract_id=1142287.
- 32 Lenzi, D., W. F. Lamb, J. Hilaire, M. Kowarsch, and J. C. Minx, 2018: Don't deploy negative
33 emissions technologies without ethical analysis. *Nature*, **561**, 303–305, doi:10.1038/d41586-
34 018-06695-5. <http://www.nature.com/articles/d41586-018-06695-5> (Accessed November 4,
35 2018).
- 36 Leon, E. G. De, and J. Pittock, 2016: Integrating climate change adaptation and climate-related
37 disaster risk-reduction policy in developing countries: A case study in the Philippines. *Clim.*
38 *Dev.*, **5529**, doi:10.1080/17565529.2016.1174659.
- 39 Levidow, L., M. Pimbert, and G. Vanloqueren, 2014: Agroecological Research: Conforming—or
40 Transforming the Dominant Agro-Food Regime? *Agroecol. Sustain. Food Syst.*, **38**, 1127–1155,
41 doi:10.1080/21683565.2014.951459.
42 <http://www.tandfonline.com/doi/abs/10.1080/21683565.2014.951459> (Accessed March 3,
43 2019).
- 44 Levis, S., A. Badger, B. Drewniak, C. Nevison, and X. Ren, 2018: CLMcrop yields and water
45 requirements: avoided impacts by choosing RCP 4.5 over 8.5. *Clim. Change*, **146**, 501–515,
46 doi:10.1007/s10584-016-1654-9.
- 47 Li, H., T. Wu, X. Wang, and Y. Qi, 2016: The Greenhouse Gas Footprint of China's Food System: An
48 Analysis of Recent Trends and Future Scenarios. *J. Ind. Ecol.*, **20**, 803–817,

- 1 doi:10.1111/jiec.12323.
- 2 Lickley, M., and S. Solomon, 2018: Drivers, timing and some impacts of global aridity change.
3 *Environ. Res. Lett.*, **13**, 104010, doi:10.1088/1748-9326/aae013. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=10/a=104010?key=crossref.83a2ca1554f5fb05ebdff31912dc2edc)
4 9326/13/i=10/a=104010?key=crossref.83a2ca1554f5fb05ebdff31912dc2edc (Accessed October
5 29, 2018).
- 6 Lin, B. B., 2011: Resilience in Agriculture through Crop Diversification: Adaptive Management for
7 Environmental Change. *Bioscience*, **61**, 183–193, doi:10.1525/bio.2011.61.3.4.
8 <https://dx.doi.org/10.1525/bio.2011.61.3.4>.
- 9 Lin, B. B., S. M. Philpott, and S. Jha, 2015: The future of urban agriculture and biodiversity-
10 ecosystem services: Challenges and next steps. *Basic Appl. Ecol.*, **16**, 189–201,
11 doi:10.1016/j.baae.2015.01.005. <http://dx.doi.org/10.1016/j.baae.2015.01.005>.
- 12 Lindblom, J., C. Lundström, M. Ljung, and A. Jonsson, 2017: Promoting sustainable intensification in
13 precision agriculture: review of decision support systems development and strategies. *Precis.*
14 *Agric.*, **18**, 309–331, doi:10.1007/s11119-016-9491-4.
- 15 Linnerooth-bayer, J., and S. Hochrainer-stigler, 2014: Financial instruments for disaster risk
16 management and climate change adaptation. *Clim. Change*, doi:10.1007/s10584-013-1035-6.
- 17 Linnerooth-Bayer, J., and S. Hochrainer-Stigler, 2015: Financial instruments for disaster risk
18 management and climate change adaptation. *Clim. Change*, **133**, 85–100, doi:10.1007/s10584-
19 013-1035-6.
- 20 Lipinski, B., C. Hanson, R. Waite, T. Searchinger, J. Lomax, and L. Kitinoja, 2013: *Reducing Food*
21 *Loss and Waste: Creating a Sustainable Food Future, Installment Two*. 1-40 pp.
22 <https://www.wri.org/publication/reducing-food-loss-and-waste> (Accessed October 4, 2018).
- 23 Lipper, L., and Coauthors, 2014: Climate-smart agriculture for food security. *Nat. Clim. Chang.*, **4**,
24 doi:10.1038/nclimate2437.
- 25 Liu, B., and Coauthors, 2016: Similar estimates of temperature impacts on global wheat yield by three
26 independent methods. *Nat. Clim. Chang.*, **6**, 1130–1136, doi:10.1038/nclimate3115.
27 <http://www.nature.com/doi/10.1038/nclimate3115>.
- 28 Liu, C., N. Hofstra, and E. Franz, 2013: Impacts of climate change on the microbial safety of pre-
29 harvest leafy green vegetables as indicated by Escherichia coli O157 and Salmonella spp. *Int. J.*
30 *Food Microbiol.*, **163**, 119–128, doi:10.1016/j.ijfoodmicro.2013.02.026.
- 31 Liu, J., and Coauthors, 2017: Challenges in operationalizing the water–energy–food nexus. *Hydrol.*
32 *Sci. J.*, **62**, 1714–1720, doi:10.1080/02626667.2017.1353695.
- 33 Liu, N., Q. Zhu, X. Zeng, B. Yang, and M. Liang, 2018a: Optimization of Processing Conditions of
34 Chinese Smoke-cured Bacon (Larou) with a New Natural Coating Solution during Storage
35 Period. **38**, 636–652.
- 36 Liu, Y., Q. Chen, Q. Ge, J. Dai, Y. Qin, L. Dai, X. Zou, and J. Chen, 2018b: Modelling the impacts of
37 climate change and crop management on phenological trends of spring and winter wheat in
38 China. *Agric. For. Meteorol.*, **248**, 518–526, doi:10.1016/J.AGRFORMET.2017.09.008.
39 <https://www.sciencedirect.com/science/article/pii/S0168192317303039?via%3Dihub> (Accessed
40 April 12, 2019).
- 41 Liu, Z., P. Yang, H. Tang, W. Wu, L. Zhang, Q. Yu, and Z. Li, 2014: Shifts in the extent and location
42 of rice cropping areas match the climate change pattern in China during 1980–2010. *Reg.*
43 *Environ. Chang.*, **15**, 919–929, doi:10.1007/s10113-014-0677-x.
- 44 Llonch, P., M. J. Haskell, R. J. Dewhurst, and S. P. Turner, 2017: Current available strategies to
45 mitigate greenhouse gas emissions in livestock systems: an animal welfare perspective. *animal*,
46 **11**, 274–284, doi:10.1017/S1751731116001440.
47 https://www.cambridge.org/core/product/identifier/S1751731116001440/type/journal_article
48 (Accessed May 25, 2018).

- 1 Lloyd, 2015: *Food system shock: The insurance impacts of acute disruption to global food supply*.
2 <https://www.lloyds.com/news-and-risk-insight/risk-reports/library/society-and-security/food->
3 [system-shock](https://www.lloyds.com/news-and-risk-insight/risk-reports/library/society-and-security/food-system-shock) (Accessed May 18, 2018).
- 4 Lobell, D., and M. Burke, 2010: *Climate change and food security: adapting agriculture to a warmer*
5 *world*. D. Lobell and M. Burke, Eds. Springer Netherlands, Dordrecht, 197 pp.
6 [http://www.worldcat.org/title/climate-change-and-food-%0A229](http://www.worldcat.org/title/climate-change-and-food-%0A229-security-adapting-agriculture-) [security-adapting-agriculture-](http://www.worldcat.org/title/climate-change-and-food-%0A229-security-adapting-agriculture-)
7 [to-a-warmer-world/oclc/630107791](http://www.worldcat.org/title/climate-change-and-food-%0A229-security-adapting-agriculture-to-a-warmer-world/oclc/630107791).
- 8 Lobell, D. B., and S. Asseng, 2017: Comparing estimates of climate change impacts from process-
9 based and statistical crop models. *Environ. Res. Lett.*, **12**, 15001, doi:10.1088/1748-
10 9326/aa518a. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=1/a=015001?key=crossref.913cd6a2c78b51ff846357868f3d2b2a)
11 [9326/12/i=1/a=015001?key=crossref.913cd6a2c78b51ff846357868f3d2b2a](http://stacks.iop.org/1748-9326/12/i=1/a=015001?key=crossref.913cd6a2c78b51ff846357868f3d2b2a) (Accessed March 7,
12 2019).
- 13 Loconto, A. M., A. Jimenez, E. Vandecastelaere, and F. Tartanac, 2018: Agroecology, local food
14 systems and their markets. *AGER J. Depopulation Rural Dev. Stud.*, **25**, 13–42,
15 doi:10.4422/AGER.2018.15. <https://hal.archives-ouvertes.fr/hal-01966250/> (Accessed April 13,
16 2019).
- 17 Loladze, I., 2014: Hidden shift of the ionome of plants exposed to elevated CO2 depletes minerals at
18 the base of human nutrition. *Elife*, **2014**, doi:10.7554/eLife.02245.
- 19 Long, T. B., V. Blok, and I. Coninx, 2016: Barriers to the adoption and diffusion of technological
20 innovations for climate-smart agriculture in Europe: Evidence from the Netherlands, France,
21 Switzerland and Italy. *J. Clean. Prod.*, **112**, 9–21, doi:10.1016/j.jclepro.2015.06.044.
22 <http://dx.doi.org/10.1016/j.jclepro.2015.06.044>.
- 23 Lopes, M. S., and Coauthors, 2015: Exploiting genetic diversity from landraces in wheat breeding for
24 adaptation to climate change. *J. Exp. Bot.*, **66**, 3477–3486, doi:10.1093/jxb/erv122.
- 25 Lopez-i-Gelats, F., 2014: Impacts of Climate Change on Food Availability: Livestock. *Global*
26 *Environmental Change*, Springer Netherlands, Dordrecht, 689–694
27 http://link.springer.com/10.1007/978-94-007-5784-4_118 (Accessed October 20, 2018).
- 28 López-i-Gelats, F., E. D. G. Fraser, J. F. Morton, and M. G. Rivera-Ferre, 2016: What drives the
29 vulnerability of pastoralists to global environmental change? A qualitative meta-analysis. *Glob.*
30 *Environ. Chang.*, **39**, 258–274, doi:10.1016/J.GLOENVCHA.2016.05.011.
31 <https://www.sciencedirect.com/science/article/abs/pii/S095937801630070X> (Accessed
32 November 3, 2018).
- 33 Lorenz, S., 2017: Adaptation planning and the use of climate change projections in local government
34 in England and Germany. *Reg. Environ. Chang.*, **17**, 425–435, doi:10.1007/s10113-016-1030-3.
- 35 Lotze-Campen, H., and Coauthors, 2014: Impacts of increased bioenergy demand on global food
36 markets: an AgMIP economic model intercomparison. *Agric. Econ.*, **45**, 103–116,
37 doi:10.1111/agec.12092. <https://doi.org/10.1111/agec.12092>.
- 38 Lozano, R., and Coauthors, 2018: *Measuring progress from 1990 to 2017 and projecting attainment*
39 *to 2030 of the health-related Sustainable Development Goals for 195 countries and territories: a*
40 *systematic analysis for the Global Burden of Disease Study 2017*. 2091–2138 pp.
41 www.thelancet.com (Accessed March 4, 2019).
- 42 Lusseau, D., and F. Mancini, 2019: Income-based variation in Sustainable Development Goal
43 interaction networks. *Nat. Sustain.*, **2**, 242–247, doi:10.1038/s41893-019-0231-4.
44 <https://doi.org/10.1038/s41893-019-0231-4>.
- 45 Lwasa, S., F. Mugagga, B. Wahab, D. Simon, J. Connors, and C. Griffith, 2014: Urban and peri-urban
46 agriculture and forestry: Transcending poverty alleviation to climate change mitigation and
47 adaptation. *Urban Clim.*, **7**, 92–106, doi:10.1016/j.uclim.2013.10.007.
48 <http://dx.doi.org/10.1016/j.uclim.2013.10.007>.

- 1 ———, ———, ———, ———, J. P. Connors, and C. Griffith, 2015: A meta-analysis of urban and peri-urban
2 agriculture and forestry in mediating climate change. *Curr. Opin. Environ. Sustain.*, **13**, 68–73,
3 doi:10.1016/j.cosust.2015.02.003. <http://dx.doi.org/10.1016/j.cosust.2015.02.003>.
- 4 Lynch, J., 2019: Availability of disaggregated greenhouse gas emissions from beef cattle production:
5 A systematic review. *Environ. Impact Assess. Rev.*, **76**, 69–78, doi:10.1016/J.EIAR.2019.02.003.
- 6 Macdiarmid, J. I., J. Kyle, G. W. Horgan, J. Loe, C. Fyfe, A. Johnstone, and G. McNeill, 2012:
7 Sustainable diets for the future: can we contribute to reducing greenhouse gas emissions by
8 eating a healthy diet? *Am. J. Clin. Nutr.*, **96**, 632–639, doi:10.3945/ajcn.112.038729.
- 9 MacDonald, G. K., K. A. Brauman, S. Sun, K. M. Carlson, E. S. Cassidy, J. S. Gerber, and P. C.
10 West, 2015: Rethinking agricultural trade relationships in an era of globalization. *Bioscience*, **65**,
11 275–289, doi:10.1093/biosci/biu225.
12 <http://bioscience.oxfordjournals.org/content/65/3/275.abstract>.
- 13 Machovina, B., K. J. Feeley, and W. J. Ripple, 2015: Biodiversity conservation: The key is reducing
14 meat consumption. *Sci. Total Environ.*, **536**, 419–431, doi:10.1016/j.scitotenv.2015.07.022.
15 <http://dx.doi.org/10.1016/j.scitotenv.2015.07.022>.
- 16 Magdoff, F., 2007: Ecological agriculture: Principles, practices, and constraints. *Renew. Agric. Food*
17 *Syst.*, **22**, 109–117, doi:10.1017/S1742170507001846.
18 http://www.journals.cambridge.org/abstract_S1742170507001846 (Accessed April 17, 2019).
- 19 Maia, S. M. F., S. M. Ogle, C. E. P. Cerri, and C. C. Cerri, 2009: Effect of grassland management on
20 soil carbon sequestration in Rondônia and Mato Grosso states, Brazil. *Geoderma*, **149**, 84–91,
21 doi:10.1016/J.GEODERMA.2008.11.023.
22 <https://www.sciencedirect.com/science/article/pii/S0016706108003406> (Accessed April 25,
23 2019).
- 24 Mairie de Paris, 2015: *Sustainable Food Plan 2015-2020*. Paris, France, 1-29 pp.
- 25 Makate, C., R. Wang, M. Makate, and N. Mango, 2016: Crop diversification and livelihoods of
26 smallholder farmers in Zimbabwe: adaptive management for environmental change.
27 *Springerplus*, **5**, 1135, doi:10.1186/s40064-016-2802-4. [https://doi.org/10.1186/s40064-016-](https://doi.org/10.1186/s40064-016-2802-4)
28 [2802-4](https://doi.org/10.1186/s40064-016-2802-4).
- 29 Malapit, H. J. L., S. Kadiyala, A. R. Quisumbing, K. Cunningham, and P. Tyagi, 2015: Women’s
30 Empowerment Mitigates the Negative Effects of Low Production Diversity on Maternal and
31 Child Nutrition in Nepal. *J. Dev. Stud.*, **51**, 1097–1123, doi:10.1080/00220388.2015.1018904.
32 <http://www.tandfonline.com/doi/full/10.1080/00220388.2015.1018904> (Accessed October 28,
33 2018).
- 34 Mallick, J., A. Rahman, and C. K. Singh, 2013: Modeling urban heat islands in heterogeneous land
35 surface and its correlation with impervious surface area by using night-time ASTER satellite
36 data in highly urbanizing city, Delhi-India. *Adv. Sp. Res.*, **52**, 639–655,
37 doi:10.1016/J.ASR.2013.04.025.
- 38 Mann, M. E., S. Rahmstorf, K. Kornhuber, B. A. Steinman, S. K. Miller, and D. Coumou, 2017:
39 Influence of Anthropogenic Climate Change on Planetary Wave Resonance and Extreme
40 Weather Events. *Sci. Rep.*, **7**, 45242, doi:10.1038/srep45242.
41 <http://www.ncbi.nlm.nih.gov/pubmed/28345645> (Accessed December 22, 2017).
- 42 Manzoor, M., S. Bibi, M. Manzoor, and R. Jabeen, 2013: Historical Analysis of Flood Information
43 and Impacts Assessment and Associated Response in Pakistan (1947-2011). *Res. J. Environ.*
44 *Earth Sci.*, **5**, 139–146.
- 45 Mao, D., F. Wang, Z. Hao, and H. Li, 2018: Credit Evaluation System Based on Blockchain for
46 Multiple Stakeholders in the Food Supply Chain. *Int. J. Environ. Res. Public Health*, **15**, 1627,
47 doi:10.3390/ijerph15081627. <http://www.ncbi.nlm.nih.gov/pubmed/30071695> (Accessed April
48 17, 2019).

- 1 Mapfumo, P., and Coauthors, 2017: Pathways to transformational change in the face of climate
2 impacts: an analytical framework. *Clim. Dev.*, **9**, 439–451,
3 doi:10.1080/17565529.2015.1040365.
4 <https://www.tandfonline.com/doi/full/10.1080/17565529.2015.1040365> (Accessed October 1,
5 2018).
- 6 Marchand, P., and Coauthors, 2016: Reserves and trade jointly determine exposure to food supply
7 shocks. *Environ. Res. Lett.*, **11**, 95009, doi:10.1088/1748-9326/11/9/095009.
8 [http://stacks.iop.org/1748-
9 9326/11/i=9/a=095009?key=crossref.15f8547c7864d2f76a79429020178381](http://stacks.iop.org/1748-9326/11/i=9/a=095009?key=crossref.15f8547c7864d2f76a79429020178381).
- 10 Marianela, F., and Coauthors, 2016: Past and present biophysical redundancy of countries as a buffer
11 to changes in food supply. *Environ. Res. Lett.*, **11**, 55008, doi:10.1088/1748-9326/11/5/055008.
12 [http://stacks.iop.org/1748-
13 9326/11/i=5/a=055008?key=crossref.5c016a9653b7213d1a23f14acd338562](http://stacks.iop.org/1748-9326/11/i=5/a=055008?key=crossref.5c016a9653b7213d1a23f14acd338562).
- 14 Marino, E., and H. Lazrus, 2015: Migration or Forced Displacement?: The Complex Choices of
15 Climate Change and Disaster Migrants in Shishmaref, Alaska and Nanumea, Tuvalu. *Hum.*
16 *Organ.*, **74**, 341–350, doi:10.17730/0018-7259-74.4.341.
17 <http://sfaajournals.net/doi/10.17730/0018-7259-74.4.341> (Accessed May 8, 2018).
- 18 Martha, G. B., E. Alves, and E. Contini, 2012: Land-saving approaches and beef production growth in
19 Brazil. *Agric. Syst.*, **110**, 173–177, doi:10.1016/J.AGSY.2012.03.001.
20 <https://www.sciencedirect.com/science/article/pii/S0308521X12000340> (Accessed June 7,
21 2018).
- 22 Marti, D. L., R. Johnson, and K. H. Mathews, 2011: *Where's the (Not) Meat? Byproducts From Beef
23 and Pork Production*.
- 24 Martin, R., B. Müller, A. Linstädter, and K. Frank, 2014: How much climate change can pastoral
25 livelihoods tolerate? Modelling rangeland use and evaluating risk. *Glob. Environ. Chang.*, **24**,
26 183–192, doi:10.1016/J.GLOENVCHA.2013.09.009.
27 <https://www.sciencedirect.com/science/article/abs/pii/S0959378013001611> (Accessed March
28 16, 2019).
- 29 —, A. Linstädter, K. Frank, and B. Müller, 2016: Livelihood security in face of drought –
30 Assessing the vulnerability of pastoral households. *Environ. Model. Softw.*, **75**, 414–423,
31 doi:10.1016/j.envsoft.2014.10.012.
- 32 Mary, S., S. Saravia-Matus, and S. Gomez y Paloma, 2018: Does nutrition-sensitive aid reduce the
33 prevalence of undernourishment? *Food Policy*, **74**, 100–116.
- 34 Masikati, P., and Coauthors, 2015: Crop-livestock intensification in the face of climate change:
35 exploring opportunities to reduce risk and increase resilience in southern Africa by using an
36 integrated multi-modeling approach. *Handbook of Climate Change and agroecosystems: The
37 Agricultural Model Intercomparison and Improvement Project Integrated Crop and Economic
38 Assessments, Part 2*, 159–198.
- 39 Mason, R., J. R. Parkins, and A. Kaler, 2017: Gendered mobilities and food security: exploring
40 possibilities for human movement within hunger prone rural Tanzania. *Agric. Human Values*,
41 **34**, 423–434, doi:10.1007/s10460-016-9723-2. [http://link.springer.com/10.1007/s10460-016-
42 9723-2](http://link.springer.com/10.1007/s10460-016-9723-2) (Accessed March 24, 2019).
- 43 Mathbout, S., J. A. Lopez-Bustins, J. Martin-Vide, J. Bech, and F. S. Rodrigo, 2018: Spatial and
44 temporal analysis of drought variability at several time scales in Syria during 1961–2012. *Atmos.*
45 *Res.*, **200**, 153–168, doi:10.1016/J.ATMOSRES.2017.09.016.
46 <https://www.sciencedirect.com/science/article/pii/S0169809517305744> (Accessed May 22,
47 2018).
- 48 Mattick, C. S., A. E. Landis, B. R. Allenby, and N. J. Genovese, 2015: Anticipatory Life Cycle
49 Analysis of In Vitro Biomass Cultivation for Cultured Meat Production in the United States.

- 1 *Environ. Sci. Technol.*, **49**, 11941–11949, doi:10.1021/acs.est.5b01614.
2 <http://www.ncbi.nlm.nih.gov/pubmed/26383898> (Accessed November 4, 2018).
- 3 Matzenberger, J., L. Kranzl, E. Tromborg, M. Junginger, V. Daioglou, C. Sheng Goh, and K.
4 Keramidias, 2015: Future perspectives of international bioenergy trade. *Renew. Sustain. Energy*
5 *Rev.*, **43**, 926–941, doi:10.1016/J.RSER.2014.10.106.
- 6 Mavromatis, T., 2015: Crop–climate relationships of cereals in Greece and the impacts of recent
7 climate trends. *Theor. Appl. Climatol.*, **120**, 417–432, doi:10.1007/s00704-014-1179-y.
8 <http://link.springer.com/10.1007/s00704-014-1179-y> (Accessed April 4, 2019).
- 9 Mayer, A., Z. Hausfather, A. D. Jones, and W. L. Silver, 2018: The potential of agricultural land
10 management to contribute to lower global surface temperatures. *Sci. Adv.*, **4**, 1–9,
11 doi:10.1126/sciadv.aag0932.
- 12 Maystadt, J.-F., and O. Ecker, 2014: Extreme weather and civil war: does drought fuel conflict in
13 Somalia through livestock price shocks? *Am. J. Agric. Econ.*, **96**, 1157–1182.
- 14 Mbow, C., O. Mertz, A. Diouf, K. Rasmussen, and A. Reenberg, 2008: The history of environmental
15 change and adaptation in eastern Saloum–Senegal—Driving forces and perceptions. *Glob.*
16 *Planet. Change*, **64**, 210–221, doi:10.1016/J.GLOPLACHA.2008.09.008.
17 <https://www.sciencedirect.com/science/article/pii/S0921818108001318> (Accessed March 9,
18 2019).
- 19 —, M. Van Noordwijk, E. Luedeling, H. Neufeldt, P. A. Minang, and G. Kowero, 2014a:
20 Agroforestry solutions to address food security and climate change challenges in Africa. *Curr.*
21 *Opin. Environ. Sustain.*, **6**, 61–67, doi:10.1016/j.cosust.2013.10.014.
22 <http://dx.doi.org/10.1016/j.cosust.2013.10.014>.
- 23 —, P. Smith, D. Skole, L. Duguma, and M. Bustamante, 2014b: Achieving mitigation and
24 adaptation to climate change through sustainable agroforestry practices in Africa. *Curr. Opin.*
25 *Environ. Sustain.*, **6**, 8–14, doi:10.1016/J.COSUST.2013.09.002.
- 26 McClanahan, T., E. H. Allison, and J. E. Cinner, 2015: Managing fisheries for human and food
27 security. *Fish Fish.*, **16**, 78–103, doi:10.1111/faf.12045. <http://doi.wiley.com/10.1111/faf.12045>
28 (Accessed April 8, 2019).
- 29 McDermid, S. P., and Coauthors, 2015: Integrated Assessments of the Impact of Climate Change on
30 Agriculture: An Overview of AgMIP Regional Research in South Asia. *Handbook of Climate*
31 *Change and agroecosystems: The Agricultural Model Intercomparison and Improvement*
32 *Project Integrated Crop and Economic Assessments, Part 2*, World Scientific, 201–217.
- 33 McGahey, D., J. Davies, N. Hagelberg, and R. Ouedraogo, 2014: *Pastoralism and the green economy:*
34 *a natural nexus? - Status, challenges and policy implications*. Nairobi, 1-72 pp.
- 35 McKinsey Global Institute, 2010: *Lions on the move: The progress and potential of African*
36 *economies* The McKinsey Global Institute. 82 pp.
- 37 McKune, S. L., and J. A. Silva, 2013: Pastoralists under Pressure: Double Exposure to Economic and
38 Environmental Change in Niger. *J. Dev. Stud.*, **49**, 1711–1727,
39 doi:10.1080/00220388.2013.822067.
40 <http://www.tandfonline.com/doi/abs/10.1080/00220388.2013.822067> (Accessed March 15,
41 2019).
- 42 —, E. C. Borresen, A. G. Young, T. D. Auria Ryley, S. L. Russo, A. Diao Camara, M. Coleman,
43 and E. P. Ryan, 2015: Climate change through a gendered lens: Examining livestock holder food
44 security. *Glob. Food Sec.*, **6**, 1–8, doi:10.1016/J.GFS.2015.05.001.
45 <https://www.sciencedirect.com/science/article/pii/S221191241500022X> (Accessed October 20,
46 2018).
- 47 McMichael, P., 2014: Historicizing food sovereignty. *J. Peasant Stud.*, **41**, 933–957.
- 48 McPhearson, T., and Coauthors, 2018: Urban Ecosystems and Biodiversity. *Rosenzweig, C., W.*

- 1 *Solecki, P. Romero-Lankao, S. Mehrotra, S. Dhakal, and S. Ali Ibrahim (eds.), Climate Change*
2 *and Cities: Second Assessment Report of the Urban Climate Change Research Network.,*
3 Cambridge University Press, New York, NY, 257–318.
- 4 Medek, D. E., J. Schwartz, and S. S. Myers, 2017: Estimated Effects of Future Atmospheric CO₂
5 Concentrations on Protein Intake and the Risk of Protein Deficiency by Country and Region.
6 *Environ. Health Perspect.*, **125**, doi:10.1289/ehp41.
- 7 Medina, A., A. Rodriguez, and N. Magan, 2014: Effect of climate change on *Aspergillus flavus* and
8 aflatoxin B₁ production. *Front. Microbiol.*, **5**, 348, doi:10.3389/fmicb.2014.00348.
9 <http://www.ncbi.nlm.nih.gov/pubmed/25101060> (Accessed January 8, 2018).
- 10 Megersa, B., A. Markemann, A. Angassa, J. O. Ogutu, H.-P. Piepho, and A. Valle Zaráte, 2014:
11 Impacts of climate change and variability on cattle production in southern Ethiopia: Perceptions
12 and empirical evidence. *Agric. Syst.*, **130**, 23–34, doi:10.1016/J.AGSY.2014.06.002.
13 <https://www.sciencedirect.com/science/article/pii/S0308521X14000742> (Accessed March 15,
14 2019).
- 15 Megido, R. C., C. Gierts, C. Blecker, Y. Brostaux, É. Haubruge, T. Alabi, and F. Francis, 2016:
16 Consumer acceptance of insect-based alternative meat products in Western countries. *Food*
17 *Qual. Prefer.*, **52**, 237–243, doi:10.1016/J.FOODQUAL.2016.05.004.
18 <https://www.sciencedirect.com/science/article/pii/S095032931630091X> (Accessed April 23,
19 2019).
- 20 van Meijl, H., and Coauthors, 2017: *Challenges of Global Agriculture in a Climate Change Context*
21 *by 2050*.
- 22 Memmott, J., P. G. Craze, N. M. Waser, and M. V Price, 2007: Global warming and the disruption of
23 plant-pollinator interactions. *Ecol. Lett.*, **10**, 710–717, doi:10.1111/j.1461-0248.2007.01061.x.
- 24 Mendelsohn, R., A. Basist, P. Kurukulasuriya, and A. Dinar, 2007: Climate and Rural Income. *Clim.*
25 *Change*, **81**, 101–118, doi:10.1007/s10584-005-9010-5.
26 <http://link.springer.com/10.1007/s10584-005-9010-5> (Accessed March 25, 2019).
- 27 Menéndez, E., C. M. Bacon, and R. Cohen, 2013: Agroecology as a Transdisciplinary, Participatory,
28 and Action-Oriented Approach. *Agroecol. Sustain. Food Syst.*, **37**, 3–18.
- 29 Menzel, A., and Coauthors, 2006: European phenological response to climate change matches the
30 warming pattern. *Glob. Chang. Biol.*, **12**, 1969–1976, doi:10.1111/j.1365-2486.2006.01193.x.
31 <https://doi.org/10.1111/j.1365-2486.2006.01193.x>.
- 32 van der Merwe, J. D., P. C. Cloete, and M. van der Hoeven, 2016: Promoting food security through
33 indigenous and traditional food crops. *Agroecol. Sustain. Food Syst.*, **40**, 830–847,
34 doi:10.1080/21683565.2016.1159642.
35 <https://www.tandfonline.com/doi/full/10.1080/21683565.2016.1159642> (Accessed October 21,
36 2018).
- 37 Meyfroidt, P., 2017: Trade - offs between environment and livelihoods : Bridging the global land use
38 and food security discussions. doi:10.1016/j.gfs.2017.08.001.
- 39 Michael, Y. G., 2017: Vulnerability and Local Innovation in Adaptation to Climate Change among
40 the Pastoralists : Harshin District , Somali Region , Ethiopia. *Environ. Manag. Sustain. Dev.*, **6**,
41 doi:10.5296/emsd.v6i2.11211.
- 42 Michalský, M., and P. S. Hooda, 2015: Greenhouse gas emissions of imported and locally produced
43 fruit and vegetable commodities: A quantitative assessment. *Environ. Sci. Policy*, **48**, 32–43,
44 doi:<https://doi.org/10.1016/j.envsci.2014.12.018>.
45 <http://www.sciencedirect.com/science/article/pii/S1462901114002469>.
- 46 Mickelbart, M. V, P. M. Hasegawa, and J. Bailey-Serres, 2015: Genetic mechanisms of abiotic stress
47 tolerance that translate to crop yield stability. *Nat. Publ. Gr.*, **16**, 237–251, doi:10.1038/nrg3901.
- 48 Van Mierlo, K., S. Rohmer, and J. C. Gerdessen, 2017: A model for composing meat replacers:

- 1 Reducing the environmental impact of our food consumption pattern while retaining its
2 nutritional value. *J. Clean. Prod.*, **165**, 930–950, doi:10.1016/j.jclepro.2017.07.098.
3 <http://dx.doi.org/10.1016/j.jclepro.2017.07.098>.
- 4 Miles, A., M. S. DeLonge, and L. Carlisle, 2017: Triggering a positive research and policy feedback
5 cycle to support a transition to agroecology and sustainable food systems. *Agroecol. Sustain.*
6 *Food Syst.*, **41**, 855–879, doi:10.1080/21683565.2017.1331179.
7 <https://www.tandfonline.com/doi/full/10.1080/21683565.2017.1331179> (Accessed October 31,
8 2018).
- 9 Milestad, R., L. Westberg, U. Geber, and J. Björklund, 2010: Enhancing Adaptive Capacity in Food
10 Systems: Learning at Farmers' Markets in Sweden. *Ecol. Soc.*, **15**, 18.
- 11 Milner, J., R. Green, A. D. Dangour, A. Haines, Z. Chalabi, J. Spadaro, A. Markandya, and P.
12 Wilkinson, 2015: Health effects of adopting low greenhouse gas emission diets in the UK. *BMJ*
13 *Open*, **5**, e007364–e007364, doi:10.1136/bmjopen-2014-007364.
14 <http://bmjopen.bmj.com/cgi/doi/10.1136/bmjopen-2014-007364>.
- 15 Minang, P. A., M. Van Noordwijk, O. E. Freeman, C. Mbow, J. de Leeuw, and D. Catacutan, 2015:
16 *Climate-smart landscapes : multifunctionality in practice*. World Agroforestry Centre, 404 pp.
17 <http://www.worldagroforestry.org/sea/Publications/files/book/BK0179-14.pdf> (Accessed May
18 20, 2018).
- 19 Minx, J. C., and Coauthors, 2018: Negative emissions - Part 1: Research landscape and synthesis.
20 *Environ. Res. Lett.*, **13**, 63001, doi:10.1088/1748-9326/aabf9b. [https://doi.org/10.1088%2F1748-](https://doi.org/10.1088%2F1748-9326%2Faabf9b)
21 [9326%2Faabf9b](https://doi.org/10.1088%2F1748-9326%2Faabf9b).
- 22 Mira de Orduña, R., 2010: Climate change associated effects on grape and wine quality and
23 production. *Food Res. Int.*, **43**, 1844–1855, doi:10.1016/J.FOODRES.2010.05.001. w (Accessed
24 April 23, 2018).
- 25 Miraglia, M., and Coauthors, 2009: Climate change and food safety: An emerging issue with special
26 focus on Europe. *Food Chem. Toxicol.*, **47**, 1009–1021, doi:10.1016/J.FCT.2009.02.005.
27 <https://www.sciencedirect.com/science/article/pii/S0278691509000714> (Accessed April 23,
28 2018).
- 29 Mishra, A. K., and S. B. Agrawal, 2014: Cultivar Specific Response of CO₂ Fertilization on Two
30 Tropical Mung Bean (*Vigna radiata* L.) Cultivars: ROS Generation, Antioxidant Status,
31 Physiology, Growth, Yield and Seed Quality. *J. Agron. Crop Sci.*, **200**, 273–289,
32 doi:10.1111/jac.12057.
- 33 Mitchell, J. P., and Coauthors, 2017: Cover cropping and no-tillage improve soil health in an arid
34 irrigated cropping system in California's San Joaquin Valley, USA. *Soil Tillage Res.*, **165**, 325–
35 335, doi:10.1016/j.still.2016.09.001. <http://dx.doi.org/10.1016/j.still.2016.09.001>.
- 36 Moat, J., J. Williams, S. Baena, T. Wilkinson, T. W. Gole, Z. K. Challa, S. Demissew, and A. P.
37 Davis, 2017: Resilience potential of the Ethiopian coffee sector under climate change. *Nat.*
38 *Plants*, **17081**, doi:10.1038/nplants.2017.81.
- 39 Mockshell, J., and J. Kamanda, 2018: Beyond the agroecological and sustainable agricultural
40 intensification debate: Is blended sustainability the way forward? *Int. J. Agric. Sustain.*, **16**, 127–
41 149.
- 42 MOFPI, 2019: Cold Chain | Ministry of Food Processing Industries | GOI.
43 <http://mofpi.nic.in/Schemes/cold-chain> (Accessed April 24, 2019).
- 44 Mogensen, L., T. L. T. Nguyen, N. T. Madsen, O. Pontoppidan, T. Preda, and J. E. Hermansen, 2016:
45 Environmental impact of beef sourced from different production systems - focus on the
46 slaughtering stage: input and output. *J. Clean. Prod.*, **133**, 284–293,
47 doi:10.1016/J.JCLEPRO.2016.05.105.
- 48 Mok, H. F., V. G. Williamson, J. R. Grove, K. Burry, S. F. Barker, and A. J. Hamilton, 2014:

- 1 Strawberry fields forever? Urban agriculture in developed countries: A review. *Agron. Sustain.*
2 *Dev.*, **34**, 21–43, doi:10.1007/s13593-013-0156-7.
- 3 Moley, K. H., and G. A. Colditz, 2016: Effects of obesity on hormonally driven cancer in women. *Sci.*
4 *Transl. Med.*, **8**, 323ps3, doi:10.1126/scitranslmed.aad8842.
5 <http://www.ncbi.nlm.nih.gov/pubmed/26819193> (Accessed April 4, 2019).
- 6 Mondal, P., and M. Basu, 2009: Adoption of precision agriculture technologies in India and in some
7 developing countries: Scope, present status and strategies. *Prog. Nat. Sci.*, **19**, 659–666,
8 doi:10.1016/J.PNSC.2008.07.020.
- 9 Moniruzzaman, S., 2015: Crop choice as climate change adaptation: Evidence from Bangladesh. *Ecol.*
10 *Econ.*, **118**, 90–98, doi:10.1016/j.ecolecon.2015.07.012.
11 <http://dx.doi.org/10.1016/j.ecolecon.2015.07.012>.
- 12 Montagnini, F., and R. Metzler, 2017: The Contribution of Agroforestry to Sustainable Development
13 Goal 2: End Hunger, Achieve Food Security and Improved Nutrition, and Promote Sustainable
14 Agriculture. Springer, Cham, 11–45 http://link.springer.com/10.1007/978-3-319-69371-2_2
15 (Accessed March 11, 2019).
- 16 Monterroso, A., C. Conde, C. Gay, D. Gómez, and J. López, 2014: Two methods to assess
17 vulnerability to climate change in the Mexican agricultural sector. *Mitig. Adapt. Strateg. Glob.*
18 *Chang.*, **19**, 445–461, doi:10.1007/s11027-012-9442-y.
19 <http://link.springer.com/10.1007/s11027-012-9442-y> (Accessed April 16, 2019).
- 20 Moore, F. C., and D. B. Lobell, 2015: The fingerprint of climate trends on European crop yields.
21 *Proc. Natl. Acad. Sci.*, 201409606.
- 22 Moragues, A., and Coauthors, 2013: *Urban Food Strategies. The rough guide to sustainable food*
23 *systems*. Document developed in the framework of the FP7 project FOODLINKS (GA No.
24 265287), 26 pp.
25 [http://www.foodlinkscommunity.net/fileadmin/documents_organicresearch/foodlinks/publicatio](http://www.foodlinkscommunity.net/fileadmin/documents_organicresearch/foodlinks/publications/Urban_food_strategies.pdf)
26 [ns/Urban_food_strategies.pdf](http://www.foodlinkscommunity.net/fileadmin/documents_organicresearch/foodlinks/publications/Urban_food_strategies.pdf).
- 27 Morrison-Whittle, P., S. A. Lee, and M. R. Goddard, 2017: Fungal communities are differentially
28 affected by conventional and biodynamic agricultural management approaches in vineyard
29 ecosystems. *Agric. Ecosyst. Environ.*, **246**, 306–313, doi:10.1016/j.agee.2017.05.022.
- 30 Morton, J., 2017: Climate change and African agriculture: unlocking the potential of research and
31 advisory services. *Making Climate Compatible Development Happen*, F. Nunan, Ed., Routledge,
32 87–113 <http://gala.gre.ac.uk/16696/>.
- 33 Morton, J. F., 2007: The impact of climate change on smallholder and subsistence agriculture. *Proc.*
34 *Natl. Acad. Sci.*, **104**, 19680–19685, doi:10.1073/pnas.0701855104.
35 <http://www.pnas.org/cgi/doi/10.1073/pnas.0701855104>.
- 36 Moses, J. A., D. S. Jayas, and K. Alagusundaram, 2015: Climate Change and its Implications on
37 Stored Food Grains. *Agric. Res.*, **4**, 21–30, doi:10.1007/s40003-015-0152-z.
38 <https://doi.org/10.1007/s40003-015-0152-z>.
- 39 Mottet, A., C. de Haan, A. Falcucci, G. Tempio, C. Opio, and P. Gerber, 2017a: Livestock: On our
40 plates or eating at our table? A new analysis of the feed/food debate. *Glob. Food Sec.*, **14**, 1–8,
41 doi:10.1016/J.GFS.2017.01.001.
42 <https://www.sciencedirect.com/science/article/pii/S2211912416300013> (Accessed March 19,
43 2018).
- 44 ———, B. Henderson, C. Opio, A. Falcucci, G. Tempio, S. Silvestri, S. Chesterman, and P. J. Gerber,
45 2017b: Climate change mitigation and productivity gains in livestock supply chains: insights
46 from regional case studies. *Reg. Environ. Chang.*, **17**, 129–141, doi:10.1007/s10113-016-0986-
47 3. <http://link.springer.com/10.1007/s10113-016-0986-3> (Accessed November 4, 2018).
- 48 Mozzer, G., 2011: Agriculture and cattle raising in the context of a low carbon economy. *Climate*

- 1 *Change in Brazil: economic, social and regulatory aspects*, R. Motta, J. Hargrave, G.
2 Luedemann, and M. Gutierrez, Eds., IPEA, Brasilia, 107–122
3 http://www.ipea.gov.br/portal/images/stories/PDFs/livros/livros/livro_climatechange_ingles.pdf.
- 4 Mueller, V., C. Gray, and K. Kosec, 2014: Heat stress increases long-term human migration in rural
5 Pakistan. *Nat. Clim. Chang.*, **4**, 182–185, doi:10.1038/nclimate2103.
- 6 Muhanji, G., R. L. Roothaert, C. Webo, and M. Stanley, 2011: African indigenous vegetable
7 enterprises and market access for small-scale farmers in East Africa. *Int. J. Agric. Sustain.*, **9**,
8 194–202, doi:10.3763/ijas.2010.0561.
9 <https://www.tandfonline.com/doi/full/10.3763/ijas.2010.0561> (Accessed October 22, 2018).
- 10 Müller, B., L. Johnson, and D. Kreuer, 2017a: Maladaptive outcomes of climate insurance in
11 agriculture. *Glob. Environ. Chang.*, **46**, 23–33, doi:10.1016/J.GLOENVCHA.2017.06.010.
12 <https://www.sciencedirect.com/science/article/pii/S0959378016304204> (Accessed March 5,
13 2019).
- 14 Müller, C., J. Elliott, and A. Levermann, 2014: Food security: Fertilizing hidden hunger. *Nat. Clim.*
15 *Chang.*, **4**, 540–541, doi:10.1038/nclimate2290.
16 <http://www.nature.com/doi/10.1038/nclimate2290> (Accessed December 6, 2017).
- 17 —, and Coauthors, 2017b: Global gridded crop model evaluation: Benchmarking, skills,
18 deficiencies and implications. *Geosci. Model Dev.*, **10**, 1403–1422, doi:10.5194/gmd-10-1403-
19 2017.
- 20 Mumtaz, Z., and P. Whiteford, 2017: Social safety nets in the development of a welfare system in
21 Pakistan : an analysis of the Benazir Income Support Programme. *Asia Pacific J. Public Adm.*,
22 **39**, 16–38, doi:10.1080/23276665.2017.1290902.
23 <http://dx.doi.org/10.1080/23276665.2017.1290902>.
- 24 Muratori, M., K. Calvin, M. Wise, P. Kyle, and J. Edmonds, 2016: Global economic consequences of
25 deploying bioenergy with carbon capture and storage (BECCS). *Environ. Res. Lett.*, **11**, 95004.
- 26 Murdiyarsa, D., and Coauthors, 2015: The potential of Indonesian mangrove forests for global climate
27 change mitigation. *Nat. Clim. Chang.*, **5**, 1089–1092, doi:10.1038/nclimate2734.
28 <http://www.nature.com/articles/nclimate2734> (Accessed April 14, 2019).
- 29 Murgueitio, E., Z. Calle, F. Uribe, A. Calle, and B. Solorio, 2011: Native trees and shrubs for the
30 productive rehabilitation of tropical cattle ranching lands. *For. Ecol. Manage.*,
31 doi:10.1016/j.foreco.2010.09.027.
- 32 Murphy, C., M. Tembo, A. Phiri, O. Yerokun, and B. Grummell, 2016: Adapting to climate change in
33 shifting landscapes of belief. *Clim. Change*, **134**, 101–114, doi:10.1007/s10584-015-1498-8.
- 34 Mutahara, M., A. Haque, M. S. A. Khan, J. F. Warner, and P. Wester, 2016: Development of a
35 sustainable livelihood security model for storm-surge hazard in the coastal areas of Bangladesh.
36 *Stoch. Environ. Res. Risk Assess.*, **30**, 1301–1315, doi:10.1007/s00477-016-1232-8.
37 <http://link.springer.com/10.1007/s00477-016-1232-8> (Accessed April 10, 2018).
- 38 Mutekwa, V. T., 2009: *Climate change impacts and adaptation in the agricultural sector: the case of*
39 *smallholder farmers in Zimbabwe*. Southern University and A & M College, 237–256 pp.
40 <https://www.cabdirect.org/cabdirect/abstract/20103307506> (Accessed April 6, 2019).
- 41 Muthayya, S., J. H. Rah, J. D. Sugimoto, F. F. Roos, K. Kraemer, and R. E. Black, 2013: The Global
42 Hidden Hunger Indices and Maps: An Advocacy Tool for Action. *PLoS One*, **8**, e67860,
43 doi:10.1371/journal.pone.0067860. <http://dx.plos.org/10.1371/journal.pone.0067860> (Accessed
44 December 1, 2017).
- 45 Muthee, K., C. Mbow, G. Macharia, and W. Leal Filho, 2017: Ecosystem-Based Adaptation (EbA) as
46 an Adaptation Strategy in Burkina Faso and Mali. 205–215
47 http://link.springer.com/10.1007/978-3-319-49520-0_13 (Accessed March 9, 2019).
- 48 Mutuo, P. K., G. Cadisch, A. Albrecht, C. A. Palm, and L. Verchot, 2005: Potential of agroforestry

- 1 for carbon sequestration and mitigation of greenhouse gas emissions from soils in the tropics.
2 *Nutr. Cycl. Agroecosystems*, **71**, 43–54, doi:10.1007/s10705-004-5285-6.
3 <http://link.springer.com/10.1007/s10705-004-5285-6> (Accessed May 25, 2018).
- 4 Myers, S. S., and Coauthors, 2014: Increasing CO₂ threatens human nutrition. *Nature*, **510**, 139–142,
5 doi:10.1038/nature13179. <http://www.nature.com/doi/10.1038/nature13179>.
- 6 —, K. R. Wessells, I. Kloog, A. Zanobetti, and J. Schwartz, 2015: Effect of increased
7 concentrations of atmospheric carbon dioxide on the global threat of zinc deficiency: A
8 modelling study. *Lancet Glob. Heal.*, **3**, e639–e645, doi:10.1016/S2214-109X(15)00093-5.
- 9 —, M. R. Smith, S. Guth, C. D. Golden, B. Vaitla, N. D. Mueller, A. D. Dangour, and P. Huybers,
10 2017: Climate Change and Global Food Systems: Potential Impacts on Food Security and
11 Undernutrition. *Annu. Rev. Public Health*, **38**, 259–277, doi:10.1146/annurev-publhealth-
12 031816-044356. [http://www.annualreviews.org/doi/10.1146/annurev-publhealth-031816-
13 044356](http://www.annualreviews.org/doi/10.1146/annurev-publhealth-031816-044356).
- 14 Næss, M. W., 2013: Climate change, risk management and the end of *Nomadic* pastoralism. *Int. J.*
15 *Sustain. Dev. World Ecol.*, **20**, 123–133, doi:10.1080/13504509.2013.779615.
16 <https://www.tandfonline.com/doi/full/10.1080/13504509.2013.779615> (Accessed March 15,
17 2019).
- 18 Nalau, J., and Coauthors, 2016: The practice of integrating adaptation and disaster risk reduction in
19 the south-west Pacific. *Clim. Dev.*, **8**, 365–375, doi:10.1080/17565529.2015.1064809.
- 20 Nandal, U., and R. L. Bhardwaj, 2014: The Role of Underutilized Fruits in Nutritional and Economic
21 Security of Tribals: A Review. *Crit. Rev. Food Sci. Nutr.*, **54**, 880–890,
22 doi:10.1080/10408398.2011.616638.
23 <http://www.tandfonline.com/doi/abs/10.1080/10408398.2011.616638> (Accessed October 21,
24 2018).
- 25 Nankishore, A., and A. D. Farrell, 2016: The response of contrasting tomato genotypes to combined
26 heat and drought stress. *J. Plant Physiol.*, **202**, 75–82, doi:10.1016/j.jplph.2016.07.006.
27 <http://www.ncbi.nlm.nih.gov/pubmed/27467552> (Accessed April 4, 2019).
- 28 Narayanan, S., and N. Gerber, 2017: Social safety nets for food and nutrition security in India. *Glob.*
29 *Food Sec.*, doi:10.1016/j.gfs.2017.05.001.
- 30 Nardone, A., B. Ronchi, N. Lacetera, M. S. Ranieri, and U. Bernabucci, 2010: Effects of climate
31 changes on animal production and sustainability of livestock systems. *Livest. Sci.*, **130**, 57–69,
32 doi:10.1016/J.LIVSCI.2010.02.011.
33 <https://www.sciencedirect.com/science/article/abs/pii/S1871141310000740> (Accessed March 5,
34 2019).
- 35 Naresh, R. K., and Coauthors, 2017: Water-Smart-Agriculture to Cope With Changing Climate in
36 Smallholders Farming Areas of Subtropical India : A Review. **5**, 400–416.
- 37 Natalini, D., G. Bravo, and A. W. Jones, 2017: Global food security and food riots--an agent-based
38 modelling approach. *Food Secur.*, 1–21.
- 39 Nautiyal, S., H. Kaechele, K. S. Rao, R. K. Maikhuri, and K. G. Saxena, 2007: Energy and economic
40 analysis of traditional versus introduced crops cultivation in the mountains of the Indian
41 Himalayas: A case study. *Energy*, **32**, 2321–2335.
- 42 Nawrotzki, R. J., F. Riosmena, and L. M. Hunter, 2013: Do rainfall deficits predict US-bound
43 migration from rural Mexico? Evidence from the Mexican census. *Popul. Res. Policy Rev.*, **32**,
44 129–158.
- 45 NCD-RisC, 2016a: Trends in adult body-mass index in 200 countries from 1975 to 2014: A pooled
46 analysis of 1698 population-based measurement studies with 19.2 million participants. *Lancet*,
47 **387**, 1377–1396, doi:10.1016/S0140-6736(16)30054-X. [http://dx.doi.org/10.1016/S0140-
48 6736\(16\)30054-X](http://dx.doi.org/10.1016/S0140-6736(16)30054-X).

- 1 —, 2016b: Worldwide trends in diabetes since 1980: a pooled analysis of 751 population-based
2 studies with 4.4 million participants. *Lancet*, **387**, 1513–1530, doi:10.1016/S0140-
3 6736(16)00618-8. <http://www.ncbi.nlm.nih.gov/pubmed/27061677> (Accessed May 9, 2018).
- 4 —, 2017a: Worldwide trends in body-mass index, underweight, overweight, and obesity from 1975
5 to 2016: a pooled analysis of 2416 population-based measurement studies in 128.9 million
6 children, adolescents, and adults. *Lancet*, **390**, 2627–2642, doi:10.1016/S0140-6736(17)32129-
7 3. <https://www.sciencedirect.com/science/article/pii/S0140673617321293> (Accessed October 8,
8 2018).
- 9 —, 2017b: Worldwide trends in blood pressure from 1975 to 2015: a pooled analysis of 1479
10 population-based measurement studies with 19.1 million participants. *Lancet (London,*
11 *England)*, **389**, 37–55, doi:10.1016/S0140-6736(16)31919-5.
12 <http://www.ncbi.nlm.nih.gov/pubmed/27863813> (Accessed May 9, 2018).
- 13 Negra, C., S. Vermeulen, L. G. Barioni, T. Mamo, P. Melville, and M. Tadesse, 2014: Brazil ,
14 Ethiopia , and New Zealand lead the way on climate-smart agriculture. 10–15,
15 doi:10.1186/s40066-014-0019-8.
- 16 Neira, D. P., and M. S. Montiel, 2013: *Agroecología y Ecofeminismo para descolonizar y*
17 *despatriarcalizar la alimentación globalizada*. 95-113 pp.
18 <https://www.upo.es/revistas/index.php/ripp/article/view/3660> (Accessed April 13, 2019).
- 19 —, X. S. Fernández, D. C. Rodríguez, M. S. Montiel, and M. D. Cabeza, 2016: Analysis of the
20 transport of imported food in Spain and its contribution to global warming. *Renew. Agric. Food*
21 *Syst.*, **31**, 37–48, doi:10.1017/S1742170514000428.
22 http://www.journals.cambridge.org/abstract_S1742170514000428 (Accessed May 31, 2018).
- 23 Nellemann, C., M. MacDevette, T. Manders, B. Eickhout, B. Svihus, A. G. Prins, and B. P.
24 Kaltenborn, 2009: *The environmental food crisis: the environment's role in averting future food*
25 *crises: a UNEP rapid response assessment*. UNEP/Earthprint, 104 pp.
26 [https://www.gwp.org/globalassets/global/toolbox/references/the-environmental-crisis.-the-](https://www.gwp.org/globalassets/global/toolbox/references/the-environmental-crisis.-the-environments-role-in-averting-future-food-crises-unep-2009.pdf)
27 [environments-role-in-averting-future-food-crises-unep-2009.pdf](https://www.gwp.org/globalassets/global/toolbox/references/the-environmental-crisis.-the-environments-role-in-averting-future-food-crises-unep-2009.pdf).
- 28 Nelson, G., M. Rosegrant, A. Palazzo, and et al., 2010: Food security, farming, and climate change to
29 2050: scenarios, results, policy options.
- 30 Nelson, G. C., and Coauthors, 2009: *Climate Change Impact on Agriculture and Costs of Adaptation*.
31 Washington, D.C., 30 pp.
32 [http://www.fao.org/fileadmin/user_upload/rome2007/docs/Impact_on_Agriculture_and_Costs_o](http://www.fao.org/fileadmin/user_upload/rome2007/docs/Impact_on_Agriculture_and_Costs_of_Adaptation.pdf)
33 [f_Adaptation.pdf](http://www.fao.org/fileadmin/user_upload/rome2007/docs/Impact_on_Agriculture_and_Costs_of_Adaptation.pdf).
- 34 Nelson, G. C., and Coauthors, 2013: Agriculture and climate change in global scenarios: why don't
35 the models agree. *Agric. Econ.*, **45**, 85–101, doi:10.1111/agec.12091.
36 <https://doi.org/10.1111/agec.12091>.
- 37 —, and Coauthors, 2018: Income Growth and Climate Change Effects on Global Nutrition Security
38 to Mid-Century. *Nat. Sustain.*, **1**, doi:10.1038/s41893-018-0192-z.
- 39 Nelson, G. C. G., and Coauthors, 2014: Climate change effects on agriculture: Economic responses to
40 biophysical shocks. *Proc. Natl. Acad. Sci.*, **111**, 3274–3279, doi:10.1073/pnas.1222465110.
41 <http://www.pnas.org/lookup/doi/10.1073/pnas.1222465110> (Accessed October 26, 2017).
- 42 Nelson, M. E., M. W. Hamm, F. B. Hu, S. A. Abrams, and T. S. Griffin, 2016: Alignment of Healthy
43 Dietary Patterns and Environmental Sustainability: A Systematic Review. *Adv. Nutr. An Int.*
44 *Rev. J.*, **7**, 1005–1025, doi:10.3945/an.116.012567.
45 <http://advances.nutrition.org/content/7/6/1005.abstract>.
- 46 Nelson, V., and T. Stathers, 2009: Resilience, power, culture, and climate: a case study from semi-arid
47 Tanzania, and new research directions. *Gend. Dev.*, **17**, 81–94,
48 doi:10.1080/13552070802696946.
49 <https://www.tandfonline.com/doi/full/10.1080/13552070802696946> (Accessed April 17, 2018).

- 1 —, K. Meadows, T. Cannon, J. Morton, and A. Martin, 2002: Uncertain predictions, invisible
2 impacts, and the need to mainstream gender in climate change adaptations. *Gend. Dev.*, **10**, 51–
3 59, doi:10.1080/13552070215911. doi:10.1080/13552070215911.
4 <https://www.tandfonline.com/doi/full/10.1080/13552070215911> (Accessed December 8, 2017).
- 5 Nemet, G. F., and Coauthors, 2018: Negative emissions - Part 3: Innovation and upscaling. *Environ.*
6 *Res. Lett.*, **13**, 63003, doi:10.1088/1748-9326/aabff4. <https://doi.org/10.1088%2F1748-9326%2Faabff4>.
7
- 8 Nepstad, D., and Coauthors, 2014: Slowing Amazon deforestation through public policy and
9 interventions in beef and soy supply chains. *Science (80-.)*, **344**, 1118–1123,
10 doi:10.1126/science.1248525. <http://science.sciencemag.org/content/344/6188/1118>.
- 11 Neufeldt, H., and Coauthors, 2013: Beyond climate-smart agriculture: toward safe operating spaces
12 for global food systems. *Agric. Food Secur.*, **2**, 12, doi:10.1186/2048-7010-2-12.
- 13 Neumayer, E., 2001: *Greening Trade and Investment: Environmental Protection Without*
14 *Protectionism*. London: Routledge, 240 pp.
15 <https://www.taylorfrancis.com/books/9781351564946>.
- 16 Newaj, R., O. P. Chaturvedi, and A. K. Handa, 2016: Recent development in agroforestry research
17 and its role in climate change adaptation and mitigation change adaptation and mitigation.
18 *Indian J. Agrofor.*, **18**, 1–9.
- 19 Newell, M. T., and R. C. Hayes, 2017: An initial investigation of forage production and feed quality
20 of perennial wheat derivatives. *Crop Pasture Sci.*, doi:10.1071/CP16405.
- 21 Newell, P., and O. Taylor, 2018: Contested landscapes: the global political economy of climate-smart
22 agriculture. *J. Peasant Stud.*, **45**, 108–129, doi:10.1080/03066150.2017.1324426.
23 <https://doi.org/10.1080/03066150.2017.1324426>.
- 24 Newman, L., C. Ling, and K. Peters, 2013: Between field and table: environmental implications of
25 local food distribution. *Int. J. Sustain. Soc.*, **5**, 11–23,
26 doi:https://doi.org/10.1504/IJSSOC.2013.050532.
27 <http://www.inderscience.com/offer.php?id=50532>.
- 28 Ng, M., and Coauthors, 2014: Global, regional, and national prevalence of overweight and obesity in
29 children and adults during 1980-2013: A systematic analysis for the Global Burden of Disease
30 Study 2013. *Lancet*, **384**, 766–781, doi:10.1016/S0140-6736(14)60460-8.
- 31 Ngadze, R. T., A. R. Linnemann, L. K. Nyanga, V. Fogliano, and R. Verkerk, 2017: Local processing
32 and nutritional composition of indigenous fruits: The case of monkey orange (*Strychnos* spp.)
33 from Southern Africa. *Food Rev. Int.*, **33**, 123–142, doi:10.1080/87559129.2016.1149862.
34 <https://www.tandfonline.com/doi/full/10.1080/87559129.2016.1149862> (Accessed October 21,
35 2018).
- 36 Nguyen, Q., J. Bow, J. Howe, S. Bratkovich, H. Groot, E. Pepke, and K. Fernholz, 2017: *Global*
37 *Production of Second Generation Biofuels: Trends and Influences*. 16 pp.
38 http://www.dovetailinc.org/report_pdfs/2017/dovetailbiofuels0117.pdf.
- 39 Nie, Z., and Coauthors, 2016: Benefits, challenges and opportunities of integrated crop-livestock
40 systems and their potential application in the high rainfall zone of southern Australia: A review.
41 *Agric. Ecosyst. Environ.*, **235**, 17–31, doi:10.1016/j.agee.2016.10.002.
42 <http://linkinghub.elsevier.com/retrieve/pii/S0167880916304947>.
- 43 Niehof, A., 2016: Food and nutrition security as gendered social practice. *Appl. Stud. Agribus.*
44 *Commer.*, **10**, 59–66, doi:10.19041/APSTRACT/2016/2-3/7.
45 http://ageconsearch.umn.edu/bitstream/250220/2/Apstract_2016_vol10_2-3_07.pdf (Accessed
46 April 25, 2018).
- 47 van Niekerk, J., and R. Wynberg, 2017: Traditional seed and exchange systems cement social
48 relations and provide a safety net: A case study from KwaZulu-Natal, South Africa. *Agroecol.*

- 1 *Sustain. Food Syst.*, 1–25, doi:10.1080/21683565.2017.1359738.
- 2 Nielsen, J. Ø., and A. Reenberg, 2010: Cultural barriers to climate change adaptation: A case study
3 from Northern Burkina Faso. *Glob. Environ. Chang.*, **20**, 142–152,
4 doi:10.1016/J.GLOENVCHA.2009.10.002.
5 <https://www.sciencedirect.com/science/article/abs/pii/S0959378009000879> (Accessed October
6 23, 2018).
- 7 Niles, M. T., and Coauthors, 2018: Climate change mitigation beyond agriculture: a review of food
8 system opportunities and implications. *Renew. Agric. Food Syst.*, **33**, 297–308,
9 doi:10.1017/S1742170518000029.
10 https://www.cambridge.org/core/product/identifier/S1742170518000029/type/journal_article.
- 11 Njiru, B. N., 2012: Climate Change, Resource Competition, and Conflict amongst Pastoral
12 Communities in Kenya. Springer, Berlin, Heidelberg, 513–527
13 http://link.springer.com/10.1007/978-3-642-28626-1_24 (Accessed March 15, 2019).
- 14 Nogueira, E. M., A. M. Yanai, S. S. de Vasconcelos, P. M. L. de Alencastro Graça, and P. M.
15 Fearnside, 2018: Brazil’s Amazonian protected areas as a bulwark against regional climate
16 change. *Reg. Environ. Chang.*, **18**, 573–579, doi:10.1007/s10113-017-1209-2.
17 <http://link.springer.com/10.1007/s10113-017-1209-2>.
- 18 Nori, M., J. Switzer, and A. Crawford, 2005: *Herding on the Brink Towards a Global Survey of*
19 *Pastoral Communities and Conflict*. Commission on Environmental, Economic and Social
20 Policy. International Institute for Sustainable Development (IISD), 1-33 pp.
21 [https://www.semanticscholar.org/paper/Herding-on-the-Brink-Towards-a-Global-Survey-of-](https://www.semanticscholar.org/paper/Herding-on-the-Brink-Towards-a-Global-Survey-of-and-Nori-Switzer/ddf99fc4a47f04356f9fdb4f8c643e5cbd4e7df8)
22 [and-Nori-Switzer/ddf99fc4a47f04356f9fdb4f8c643e5cbd4e7df8](https://www.semanticscholar.org/paper/Herding-on-the-Brink-Towards-a-Global-Survey-of-and-Nori-Switzer/ddf99fc4a47f04356f9fdb4f8c643e5cbd4e7df8) (Accessed November 4, 2018).
- 23 NRC, 1981: *Effect of environment on nutrient requirements of domestic animals. National Research*
24 *Council (U.S.). Committee on Animal Nutrition. Subcommittee on Environmental Stress.*
25 National Academy Press, 152 pp.
26 [https://books.google.es/books?hl=en&lr=&id=n1V_Nnz_NWEC&oi=fnd&pg=PT10&ots=2uZs](https://books.google.es/books?hl=en&lr=&id=n1V_Nnz_NWEC&oi=fnd&pg=PT10&ots=2uZs_C25tS&sig=x20bWkB0grcpVaZUopx-bVTOsTY&redir_esc=y#v=onepage&q&f=false)
27 [_C25tS&sig=x20bWkB0grcpVaZUopx-bVTOsTY&redir_esc=y#v=onepage&q&f=false](https://books.google.es/books?hl=en&lr=&id=n1V_Nnz_NWEC&oi=fnd&pg=PT10&ots=2uZs_C25tS&sig=x20bWkB0grcpVaZUopx-bVTOsTY&redir_esc=y#v=onepage&q&f=false)
28 (Accessed March 16, 2019).
- 29 NRDC, 2017: *Less beef, less carbon: Americans shrink their diet-related carbon footprint by 10*
30 *percent between 2005 and 2014.* 8 pp.
- 31 Nurse, L. A., R. F. McLean, J. Agard, L. P. Briguglio, V. Duvat-Magnan, N. Pelesikoti, E. Tompkins,
32 and A. Webb, 2014: *Small Islands. In: Climate Change 2014: Impacts, Adaptation, and*
33 *Vulnerability. Part B: Regional Aspects. Contribution of Working Group II to the Fifth*
34 *Assessment Report of the Intergovernmental Panel on Climate Change [Barros, V.R., C.B.*
35 *Field, D.J. D. Cambridge, UK and New York, USA, 1613-1654 pp.*
- 36 Nyong, A., F. Adesina, and B. Osman Elasha, 2007: The value of indigenous knowledge in climate
37 change mitigation and adaptation strategies in the African Sahel. *Mitig. Adapt. Strateg. Glob.*
38 *Chang.*, **12**, 787–797, doi:10.1007/s11027-007-9099-0.
39 <http://link.springer.com/10.1007/s11027-007-9099-0> (Accessed March 9, 2019).
- 40 O’Brien, K. L., and J. Wolf, 2010: A values-based approach to vulnerability and adaptation to climate
41 change. *Wiley Interdiscip. Rev. Clim. Chang.*, **1**, 232–242, doi:10.1002/wcc.30.
42 <http://doi.wiley.com/10.1002/wcc.30> (Accessed October 23, 2018).
- 43 O’Mara, F. P., 2012: The role of grasslands in food security and climate change. *Ann. Bot.*, **110**,
44 1263–1270, doi:10.1093/aob/mcs209. <http://www.ncbi.nlm.nih.gov/pubmed/23002270>
45 (Accessed April 17, 2019).
- 46 Obersteiner, M., and Coauthors, 2016: Assessing the land resource–food price nexus of the
47 Sustainable Development Goals. *Sci. Adv.*, **2**, e1501499, doi:10.1126/sciadv.1501499.
48 <http://advances.sciencemag.org/lookup/doi/10.1126/sciadv.1501499> (Accessed November 8,
49 2018).

- 1 OECD, 2017: *International Migration Outlook 2017*. OECD Publishing, Paris, 366 pp.
2 [http://www.oecd-ilibrary.org/social-issues-migration-health/international-migration-outlook-](http://www.oecd-ilibrary.org/social-issues-migration-health/international-migration-outlook-2017_migr_outlook-2017-en)
3 [2017_migr_outlook-2017-en](http://www.oecd-ilibrary.org/social-issues-migration-health/international-migration-outlook-2017_migr_outlook-2017-en).
- 4 —, and FAO, 2015: *OECD-FAO Agricultural Outlook 2015*. Paris, 148 pp pp.
5 <http://www.fao.org/3/a-i4738e.pdf>.
- 6 Ojha, H. R., S. Ghimire, A. Pain, A. Nightingale, B. Khatri, and H. Dhungana, 2015: Policy without
7 politics: technocratic control of climate change adaptation policy making in Nepal.
8 doi:10.1080/14693062.2014.1003775.
- 9 Oliveira, G., and S. Hecht, 2016: Sacred groves, sacrifice zones and soy production: globalization,
10 intensification and neo-nature in South America. *J. Peasant Stud.*, **43**, 251–285,
11 doi:10.1080/03066150.2016.1146705.
12 <http://www.tandfonline.com/doi/full/10.1080/03066150.2016.1146705>.
- 13 de Oliveira Silva, R., L. G. Barioni, J. A. J. Hall, M. Folegatti Matsuura, T. Zanett Albertini, F. A.
14 Fernandes, and D. Moran, 2016: Increasing beef production could lower greenhouse gas
15 emissions in Brazil if decoupled from deforestation. *Nat. Clim. Chang.*, **6**, 493–497,
16 doi:10.1038/nclimate2916. <http://www.nature.com/doi/10.1038/nclimate2916>.
- 17 Oliver, T. H., and Coauthors, 2018: Overcoming undesirable resilience in the global food system.
18 *Glob. Sustain.*, **1**.
- 19 Omolo, N., 2011: Gender and climate change-induced conflict in pastoral communities: Case study of
20 Turkana in northwestern Kenya. *African J. Confl. Resolut.*, **10**, doi:10.4314/ajcr.v10i2.63312.
21 <http://www.ajol.info/index.php/ajcr/article/view/63312> (Accessed March 15, 2019).
- 22 Omotilewa, O. J., J. Ricker-Gilbert, J. H. Ainembabazi, and G. E. Shively, 2018: Does improved
23 storage technology promote modern input use and food security? Evidence from a randomized
24 trial in Uganda. *J. Dev. Econ.*, **135**, 176–198, doi:10.1016/j.jdeveco.2018.07.006.
25 <https://doi.org/10.1016/j.jdeveco.2018.07.006>.
- 26 Ongoro, E., and W. Ogara, 2012: Impact of climate change and gender roles in community
27 adaptation: A case study of pastoralists in Samburu East District, Kenya. *Int. J. Biodivers.*
28 *Conserv.*, **42**, 78–89.
29 [https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_r](https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_roles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Kenya)
30 [oles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Kenya](https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_roles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Kenya)
31 [a](https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_roles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Kenya) (Accessed March 24, 2019).
- 32 Onyeneke, R. U., 2018: *International journal of biosciences, agriculture and technology : IJBSAT*. A.
33 Parbhu Britto, 1-7 pp. <https://www.cabdirect.org/cabdirect/abstract/20183191054> (Accessed
34 November 4, 2018).
- 35 —, C. O. Igberu, C. O. Uwadoka, and J. O. Aligbe, 2018: Status of climate-smart agriculture in
36 southeast Nigeria. *GeoJournal*, **83**, 333–346, doi:10.1007/s10708-017-9773-z.
37 <http://link.springer.com/10.1007/s10708-017-9773-z> (Accessed April 4, 2019).
- 38 Ortega, F. B., C. J. Lavie, and S. N. Blair, 2016: Obesity and Cardiovascular Disease. *Circ. Res.*, **118**,
39 1752–1770, doi:10.1161/CIRCRESAHA.115.306883.
40 <https://www.ahajournals.org/doi/10.1161/CIRCRESAHA.115.306883> (Accessed April 4, 2019).
- 41 Ouma, E. A., G. A. Obare, and S. J. Staal, 2003: CATTLE AS ASSETS: ASSESSMENT OF NON-
42 MARKET BENEFITS FROM CATTLE IN SMALLHOLDER KENYAN CROP-LIVESTOCK
43 SYSTEMS.
- 44 Ozaki, M., 2016: *Disaster Risk Financing in Bangladesh Disaster Risk Financing in Bangladesh*.
- 45 Padgham, J., J. Jabbour, and K. Dietrich, 2015: Managing change and building resilience: A multi-
46 stressor analysis of urban and peri-urban agriculture in Africa and Asia. *Urban Clim.*, **12**, 183–
47 204, doi:10.1016/j.uclim.2015.04.003. <http://dx.doi.org/10.1016/j.uclim.2015.04.003>.
- 48 Padulosi, S., J. Thompson, and P. Ruderbjerg, 2013: *Fighting poverty, hunger and malnutrition with*

- 1 *neglected and underutilized species (NUS): needs, challenges and the way forward. Bioversity*
2 *International, Rome.*
- 3 Palm, C., H. Blanco-canqui, F. Declerck, L. Gatere, and P. Grace, 2013: Conservation agriculture and
4 ecosystem services: An overview. *"Agriculture, Ecosyst. Environ.,*
5 *doi:10.1016/j.agee.2013.10.010.*
- 6 Palm, C. A., and Coauthors, 2010: Identifying potential synergies and trade-offs for meeting food
7 security and climate change objectives in sub-Saharan Africa. *Proc. Natl Acad. Sci.,* **107**, 19661,
8 *doi:10.1073/pnas.0912248107.*
- 9 Palmer, M. A., J. Liu, J. H. Matthews, M. Mumba, and P. D'Odorico, 2015: Manage water in a
10 green way. *Science (80-.),* **349**, 584 LP-585, *doi:10.1126/science.aac7778.*
11 [http://science.sciencemag.org/content/349/6248/584.2.abstract.](http://science.sciencemag.org/content/349/6248/584.2.abstract)
- 12 Pangga, I. B., J. Hanan, and S. Chakraborty, 2011: Pathogen dynamics in a crop canopy and their
13 evolution under changing climate. *Plant Pathol.,* **60**, 70–81, *doi:10.1111/j.1365-*
14 *3059.2010.02408.x.*
- 15 Pant, L. P., B. Adhikari, and K. K. Bhattarai, 2015: Adaptive transition for transformations to
16 sustainability in developing countries. *Curr. Opin. Environ. Sustain.,* **14**, 206–212.
- 17 Papageorgiou, M., 2017: iMedPub Journals Genomics-Centric Approach : An Insight into the Role of
18 Genomics in Assisting GM-Rice Varieties With-in a Paradigm of Future Climate-Change
19 Mitigation , Food Security , and GM Regulation Keywords : Genomic Tools Used in Rice
20 Improvement. *J. Plant Sci. Agric. Res.,* **1**, 1–10.
- 21 Papargyropoulou, E., R. Lozano, J. K. Steinberger, N. Wright, and Z. bin Ujang, 2014: The food
22 waste hierarchy as a framework for the management of food surplus and food waste. *J. Clean.*
23 *Prod.,* **76**, 106–115, *doi:10.1016/j.jclepro.2014.04.020.*
24 <https://linkinghub.elsevier.com/retrieve/pii/S0959652614003680> (Accessed March 17, 2019).
- 25 Pape, R., and J. Löffler, 2012: Climate Change, Land Use Conflicts, Predation and Ecological
26 Degradation as Challenges for Reindeer Husbandry in Northern Europe: What do We Really
27 Know After Half a Century of Research? *Ambio,* **41**, 421–434, *doi:10.1007/s13280-012-0257-6.*
28 <http://link.springer.com/10.1007/s13280-012-0257-6> (Accessed March 16, 2019).
- 29 Parikh, J., 2009: *Towards a gender-sensitive agenda for energy, environment and climate change.*
30 Division for the Advancement of Women, United Nations, Geneva, 7 pp.
31 [http://www.un.org/womenwatch/daw/egm/impact_bdpfa/EP10 - Parikh_final.after track.pdf.](http://www.un.org/womenwatch/daw/egm/impact_bdpfa/EP10 - Parikh_final.after track.pdf)
- 32 Paris, T., and M. F. Rola-Rubzen, 2018: *Gender dimension of climate change research in agriculture:*
33 *Case studies in Southeast Asia.* CGIAR research program on Climate Change, Agriculture and
34 Food Security, Wageningen, 216 pp. <https://ccafs.cgiar.org/fr/node/56839#.XJKHPShKg2w>
35 (Accessed March 20, 2019).
- 36 Park, S. E., N. A. Marshall, E. Jakku, A. M. Dowd, S. M. Howden, E. Mendham, and A. Fleming,
37 2012: Informing adaptation responses to climate change through theories of transformation.
38 *Glob. Environ. Chang.,* **22**, 115–126, *doi:10.1016/J.GLOENVCHA.2011.10.003.*
39 <https://www.sciencedirect.com/science/article/pii/S0959378011001555> (Accessed April 23,
40 2019).
- 41 Parkes, B., D. Defrance, B. Sultan, P. Ciais, and X. Wang, 2018: Projected changes in crop yield
42 mean and variability over West Africa in a world 1.5 K warmer than the pre-industrial. 1–24,
43 *doi:10.5194/esd-2017-66.*
- 44 Pascucci, S., C. Cicatiello, S. Franco, B. Pancino, D. Marinov, and M. Davide, 2011: *The*
45 *international food and agribusiness management review.* International Food and Agribusiness
46 Management Association, 2003-, 105-126 pp. <https://library.wur.nl/WebQuery/wurpubs/414639>
47 (Accessed March 25, 2019).
- 48 Pasini, A., G. Mastrojeni, and F. N. Tubiello, 2018: Climate actions in a changing world. *Anthr. Rev.,*

- 1 rangelands – Insights from a trait-based dynamic vegetation model. *Ecol. Modell.*, **395**, 11–22,
2 doi:10.1016/j.ecolmodel.2018.12.013.
- 3 Pham, T. P. T., R. Kaushik, G. K. Parshetti, R. Mahmood, and R. Balasubramanian, 2015: Food
4 waste-to-energy conversion technologies: Current status and future directions. *Waste Manag.*,
5 **38**, 399–408, doi:10.1016/J.WASMAN.2014.12.004.
6 <https://www.sciencedirect.com/science/article/pii/S0956053X14005819?via%3Dihub> (Accessed
7 March 17, 2019).
- 8 Pikaar, I., S. Matassa, K. Rabaey, B. Laycock, N. Boon, and W. Verstraete, 2018: The Urgent Need to
9 Re-engineer Nitrogen-Efficient Food Production for the Planet. *Managing Water, Soil and*
10 *Waste Resources to Achieve Sustainable Development Goals*, Springer International Publishing,
11 Cham, 35–69 http://link.springer.com/10.1007/978-3-319-75163-4_3 (Accessed May 31, 2018).
- 12 Pimbert, M., 2015: Agroecology as an Alternative Vision to Conventional Development and Climate-
13 smart Agriculture. *Development*, **58**, 286–298, doi:10.1057/s41301-016-0013-5.
14 <http://link.springer.com/10.1057/s41301-016-0013-5> (Accessed March 3, 2019).
- 15 —, and S. Lemke, 2018: Using agroecology to enhance dietary diversity. *UNSCN News*, **43**, 33–42.
16 <https://pureportal.coventry.ac.uk/en/publications/using-agroecology-to-enhance-dietary-diversity>
17 (Accessed November 4, 2018).
- 18 Pinke, Z., and G. L. Lövei, 2017: Increasing temperature cuts back crop yields in Hungary over the
19 last 90 years. *Glob. Chang. Biol.*, **23**, 5426–5435, doi:10.1111/gcb.13808.
20 <http://www.ncbi.nlm.nih.gov/pubmed/28699259> (Accessed April 4, 2019).
- 21 Place, S. E., and F. M. Mitloehner, 2012: Beef production in balance: Considerations for life cycle
22 analyses. *Meat Sci.*, **92**, 179–181, doi:10.1016/J.MEATSCI.2012.04.013.
- 23 Ponnusamy, P., and Coauthors, 2015: Integrated Assessment of Climate Change Impacts on Maize
24 Farms and Farm Household Incomes in South India: A Case Study from Tamil Nadu. 9;
25 Chapter.
- 26 Poore, J., and T. Nemecek, 2018: Reducing food’s environmental impacts through producers and
27 consumers. *Science*, **360**, 987–992, doi:10.1126/science.aag0216.
28 <http://www.ncbi.nlm.nih.gov/pubmed/29853680> (Accessed November 4, 2018).
- 29 Popp, A., and Coauthors, 2017: Land-use futures in the shared socio-economic pathways. *Glob.*
30 *Environ. Chang.*, **42**, 331–345, doi:10.1016/j.gloenvcha.2016.10.002.
31 <http://dx.doi.org/10.1016/j.gloenvcha.2016.10.002>.
- 32 Popp, J., Z. Lakner, M. Harangi-Rákos, and M. Fári, 2014: The effect of bioenergy expansion: Food,
33 energy, and environment. *Renew. Sustain. Energy Rev.*, **32**, 559–578,
34 doi:10.1016/J.RSER.2014.01.056.
35 <https://www.sciencedirect.com/science/article/pii/S1364032114000677?via%3Dihub> (Accessed
36 November 5, 2018).
- 37 Porter, J. R., L. Xie, A. J. Challinor, K. Cochrane, S. M. Howden, M. M. Iqbal, D. B. Lobell, and M.
38 I. Travasso, 2014: Food security and food production systems. *Clim. Chang. 2014 Impacts,*
39 *Adapt. Vulnerability. Part A Glob. Sect. Asp. Contrib. Work. Gr. II to Fifth Assess. Rep.*
40 *Intergov. Panel Clim. Chang.*, **2**, 485–533, doi:10.1111/j.1728-4457.2009.00312.x.
- 41 —, M. Howden, and P. Smith, 2017: Considering agriculture in IPCC assessments. *Nat. Clim.*
42 *Chang.*, **7**, 680–683, doi:10.1038/nclimate3404.
43 <http://www.nature.com/doi/10.1038/nclimate3404> (Accessed October 1, 2018).
- 44 Porter, S. D., D. S. Reay, P. Higgins, and E. Bomberg, 2016: A half-century of production-phase
45 greenhouse gas emissions from food loss & waste in the global food supply chain. *Sci. Total*
46 *Environ.*, **571**, 721–729, doi:10.1016/j.scitotenv.2016.07.041.
47 <http://dx.doi.org/10.1016/j.scitotenv.2016.07.041>.
- 48 Potopová, V., P. Zahradníček, P. Štěpánek, L. Türkott, A. Farda, and J. Soukup, 2017: The impacts of

- 1 key adverse weather events on the field-grown vegetable yield variability in the Czech Republic
2 from 1961 to 2014. *Int. J. Climatol.*, **37**, 1648–1664, doi:10.1002/joc.4807.
3 <http://doi.wiley.com/10.1002/joc.4807> (Accessed November 4, 2018).
- 4 Powlson, D. S., C. M. Stirling, M. L. Jat, B. G. Gerard, C. A. Palm, P. A. Sanchez, and K. G.
5 Cassman, 2014: Limited potential of no-till agriculture for climate change mitigation/change
6 mitigation. *Nat. Clim. Chang.*, **4**, 678–683, doi:10.1038/NCLIMATE2292.
- 7 Powlson, D. S., C. M. Stirling, C. Thierfelder, R. P. White, and M. L. Jat, 2016: Does conservation
8 agriculture deliver climate change mitigation through soil carbon sequestration in tropical agro-
9 ecosystems? *Agric. Ecosyst. Environ.*, **220**, 164–174, doi:10.1016/j.agee.2016.01.005.
- 10 Pradhan, P., 2019: Antagonists to meeting the 2030 Agenda. *Nat. Sustain.*, **2**, 171–172,
11 doi:10.1038/s41893-019-0248-8. <https://doi.org/10.1038/s41893-019-0248-8>.
- 12 —, M. K. B. Lüdeke, D. E. Reusser, and J. P. Kropp, 2013a: Embodied crop calories in animal
13 products. *Environ. Res. Lett.*, **8**, 44044, doi:<http://dx.doi.org/10.1088/1748-9326/8/4/044044>.
14 <http://stacks.iop.org/1748-9326/8/i=4/a=044044%5Cnhttp://iopscience.iop.org/1748->
15 [9326/8/4/044044/pdf/1748-9326_8_4_044044.pdf](http://stacks.iop.org/1748-9326/8/4/044044/pdf/1748-9326_8_4_044044.pdf).
- 16 —, D. E. Reusser, and J. P. Kropp, 2013b: Embodied Greenhouse Gas Emissions in Diets. *PLoS*
17 *One*, **8**, 1–8, doi:10.1371/journal.pone.0062228.
- 18 —, D. E. Reusser, and J. P. Kropp, 2014: Food Self-Sufficiency across Scales: How Local Can We
19 Go? **15**, 9779.
- 20 —, G. Fischer, H. Van Velthuis, D. E. Reusser, and J. P. Kropp, 2015: Closing yield gaps: How
21 sustainable can we be? *PLoS One*, **10**, 1–18, doi:10.1371/journal.pone.0129487.
22 <http://dx.doi.org/10.1371/journal.pone.0129487>.
- 23 —, L. Costa, D. Rybski, W. Lucht, and J. P. Kropp, 2017: A Systematic Study of Sustainable
24 Development Goal (SDG) Interactions. *Earth's Futur.*, **5**, 1169–1179,
25 doi:10.1002/2017EF000632.
- 26 Pratibha, G., and Coauthors, 2016: Net global warming potential and greenhouse gas intensity of
27 conventional and conservation agriculture system in rainfed semi arid tropics of India. *Atmos.*
28 *Environ.*, **145**, 239–250, doi:10.1016/J.ATMOSENV.2016.09.039.
29 <https://www.sciencedirect.com/science/article/pii/S1352231016307397> (Accessed March 9,
30 2019).
- 31 Pretty, J., and Z. P. Bharucha, 2014: Sustainable intensification in agricultural systems. *Ann. Bot.*,
32 **114**, 1571–1596, doi:10.1093/aob/mcu205.
- 33 —, and —, 2015: Integrated Pest Management for Sustainable Intensification of Agriculture in
34 Asia and Africa. *Insects*, **6**, 152–182, doi:10.3390/insects6010152.
35 <https://www.ncbi.nlm.nih.gov/pubmed/26463073>.
- 36 —, and Coauthors, 2018: Global assessment of agricultural system redesign for sustainable
37 intensification. *Nat. Sustain.*, doi:10.1038/s41893-018-0114-0.
- 38 Pugh, T. A. M., C. Müller, J. Elliott, D. Deryng, C. Folberth, S. Olin, E. Schmid, and A. Arneth,
39 2016: Climate analogues suggest limited potential for intensification of production on current
40 croplands under climate change. *Nat. Commun.*, **7**, 1–8, doi:10.1038/ncomms12608.
- 41 Puma, M. J., S. Bose, S. Y. Chon, and B. I. Cook, 2015: Assessing the evolving fragility of the global
42 food system. *Environ. Res. Lett.*, **10**, 24007, doi:10.1088/1748-9326/10/2/024007.
43 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/10/i=2/a=024007?key=crossref.d345e1eb75f8addcbd02eefe7ee6c482)
44 [9326/10/i=2/a=024007?key=crossref.d345e1eb75f8addcbd02eefe7ee6c482](http://stacks.iop.org/1748-9326/10/i=2/a=024007?key=crossref.d345e1eb75f8addcbd02eefe7ee6c482).
- 45 Le Quéré, C., and Coauthors, 2018: Global Carbon Budget 2018. *Earth Syst. Sci. Data*, **10**, 2141–
46 2194, doi:10.5194/essd-10-2141-2018. <https://www.earth-syst-sci-data.net/10/2141/2018/>
47 (Accessed April 16, 2019).

- 1 Rabbani, M. G., Z. M. Khan, and M. H. Tuhin, 2015: *Climate change and migration in Bangladesh: A*
2 *gender perspective*. Bangladesh, 49 pp. [http://asiapacific.unwomen.org/en/digital-](http://asiapacific.unwomen.org/en/digital-library/publications/2016/01/climate-change-and-migration-in-bangladesh)
3 [library/publications/2016/01/climate-change-and-migration-in-bangladesh](http://asiapacific.unwomen.org/en/digital-library/publications/2016/01/climate-change-and-migration-in-bangladesh).
- 4 Rahman, M. R., and S. H. Bulbul, 2015: Adoption of Water Saving Irrigation Techniques for
5 Sustainable Rice Production in Bangladesh. *Environ. Ecol. Res.*, **3**, 1–8,
6 doi:10.13189/eer.2015.030101.
- 7 Rajkumar, M., M. N. V Prasad, S. Swaminathan, and H. Freitas, 2013: Climate change driven plant-
8 metal-microbe interactions. *Environ. Int.*, **53**, 74–86, doi:10.1016/j.envint.2012.12.009.
- 9 Raleigh, C., and D. Kniveton, 2012: Come rain or shine: An analysis of conflict and climate
10 variability in East Africa. *J. Peace Res.*, **49**, 51–64, doi:10.1177/0022343311427754.
11 <http://journals.sagepub.com/doi/10.1177/0022343311427754> (Accessed March 15, 2019).
- 12 Ramachandran Nair, P. K., B. Mohan Kumar, and V. D. Nair, 2009: Agroforestry as a strategy for
13 carbon sequestration. *J. Plant Nutr. Soil Sci.*, **172**, 10–23, doi:10.1002/jpln.200800030.
14 <http://doi.wiley.com/10.1002/jpln.200800030> (Accessed May 25, 2018).
- 15 Rao, C. A. R., and Coauthors, 2016: *A district level assessment of vulnerability of Indian agriculture*
16 *to climate change*. http://eprints.icrisat.ac.in/14416/1/Current_Sci_110_10_1939-1946_2016.pdf
17 (Accessed February 25, 2019).
- 18 Rao, K. P. C., G. Sridhar, R. M. Mulwa, M. N. Kilavi, A. Esilaba, I. N. Athanasiadis, and R. O.
19 Valdivia, 2015: Impacts of climate variability and change on agricultural systems in East Africa.
20 *Handbook of climate change and agroecosystems: The Agricultural Model Intercomparison and*
21 *Improvement Project Integrated Crop and Economic Assessments, Part 2*, World Scientific, 75–
22 124.
- 23 Rao, K. V. V, A. K. Sikka, A. Islam, and K. V. V Rao, 2017: Climate-Smart Land and Water
24 Management for Sustainable Agriculture. *Irrig. Drain.*, doi:10.1002/ird.2162.
25 <http://doi.wiley.com/10.1002/ird.2162>.
- 26 Rao, N., 2005: Gender Equality, Land Rights and Household Food Security: Discussion of Rice
27 Farming Systems. *Econ. Polit. Wkly.*, **40**, 2513–2521, doi:10.2307/4416780.
28 <https://www.jstor.org/stable/4416780> (Accessed December 8, 2017).
- 29 Rasul, G., 2010: The role of the Himalayan mountain systems in food security and agricultural
30 sustainability in South Asia. *Int. J. Rural Manag.*, **6**, 95–116.
- 31 Rasul, G., A. Saboor, P. C. Tiwari, A. Hussain, N. Ghosh, and G. B. Chettri, 2019: Food and
32 Nutritional Security in the Hindu Kush Himalaya: Unique Challenges and Niche Opportunities.
33 *The Hindu Kush Himalaya Assessment: mountains, climate change, sustainability and people.*,
34 P. Wester, A. Mishra, A. Mukherji, and A.B. Shrestha, Eds., SPRINGER, Dordrecht
35 <https://www.springer.com/us/book/9783319922874> (Accessed October 29, 2018).
- 36 Rattanachot, W., Y. Wang, D. Chong, and S. Suwansawas, 2015: Adaptation strategies of transport
37 infrastructures to global climate change. *Transp. Policy*, **41**, 159–166,
38 doi:10.1016/j.tranpol.2015.03.001. <http://dx.doi.org/10.1016/j.tranpol.2015.03.001>.
- 39 Ravi, S., D. D. Breshears, T. E. Huxman, and P. D’Odorico, 2010: Land degradation in drylands:
40 Interactions among hydrologic–aeolian erosion and vegetation dynamics. *Geomorphology*, **116**,
41 236–245, doi:10.1016/J.GEOMORPH.2009.11.023.
42 <https://www.sciencedirect.com/science/article/pii/S0169555X09005108> (Accessed May 25,
43 2018).
- 44 Reay, D. S., E. A. Davidson, K. A. Smith, P. Smith, J. M. Melillo, F. Dentener, and P. J. Crutzen,
45 2012: Global agriculture and nitrous oxide emissions. *Nat. Clim. Chang.*, **2**, 410–416,
46 doi:10.1038/nclimate1458. <http://dx.doi.org/10.1038/nclimate1458>.
- 47 Red Cross, 2011: West and Central Africa: Early Warning / Early Action Red Cross and Red
48 Crescent action.

- 1 Remling, E., and J. Veitayaki, 2016: Community-based action in Fiji's Gau Island: a model for the
2 Pacific? *Int. J. Clim. Chang. Strateg. Manag.*, **8**, 375–398, doi:10.1108/IJCCSM-07-2015-0101.
3 <http://www.emeraldinsight.com/doi/10.1108/IJCCSM-07-2015-0101> (Accessed October 23,
4 2018).
- 5 Ren, X., M. Weitzel, B. C. O'Neill, P. Lawrence, P. Meiyappan, S. Levis, E. J. Balistreri, and M.
6 Dalton, 2018: Avoided economic impacts of climate change on agriculture: integrating a land
7 surface model (CLM) with a global economic model (iPETS). *Clim. Change*, **146**, 517–531,
8 doi:10.1007/s10584-016-1791-1. <http://link.springer.com/10.1007/s10584-016-1791-1>
9 (Accessed April 16, 2019).
- 10 Reuter, T., 2018: Understanding Food System Resilience in Bali, Indonesia: A Moral Economy
11 Approach. *Cult. Agric. Food Environ.*, **0**, doi:10.1111/cuag.12135.
12 <https://doi.org/10.1111/cuag.12135>.
- 13 Reuter, T. A., 2015: The Struggle for Food Sovereignty: A Global Perspective. *Averting a Global*
14 *Environmental Collapse*, 127–147.
- 15 Revi, A., D. E. Satterthwaite, F. Aragón-Durand, J. Corfee-Morlot, R. B. R. Kiunsi, M. Pelling, D. C.
16 Roberts, and W. Solecki, 2014: Urban areas Climate Change 2014: Impacts, Adaptation, and
17 Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the
18 Fifth Assessment Report of the Intergovernmental Panel on Climate Change ed CB Field et al.
19 *Field, CB, Barros, VR, Dokken, DJ al*, 535–612.
- 20 Reyer, C. P., and Coauthors, 2017a: Climate change impacts in Latin America and the Caribbean and
21 their implications for development. *Reg. Environ. Chang.*, **17**, 1601–1621, doi:10.1007/s10113-
22 015-0854-6. <https://link.springer.com/content/pdf/10.1007%2Fs10113-015-0854-6.pdf>
23 (Accessed February 27, 2019).
- 24 Reyer, C. P. O., & Kanta, K. Rigaud, E. Fernandes, W. Hare, O. Serdeczny, and H. J. Schellnhuber,
25 2017b: Turn down the heat: regional climate change impacts on development. *Reg. Environ.*
26 *Chang.*, **17**, 1563–1568, doi:10.1007/s10113-017-1187-4.
27 <https://link.springer.com/content/pdf/10.1007%2Fs10113-017-1187-4.pdf> (Accessed February
28 27, 2019).
- 29 Riahi, K., and Coauthors, 2017: The Shared Socioeconomic Pathways and their energy, land use, and
30 greenhouse gas emissions implications: An overview. *Glob. Environ. Chang.*, **42**, 153–168.
- 31 Richards, M., and Coauthors, 2015: *How countries plan to address agricultural adaptation and*
32 *mitigation An analysis of Intended Nationally Determined Contributions*. 8 pp.
33 [https://cgspace.cgiar.org/bitstream/handle/10568/69115/CCAFS_INDC_info_note-](https://cgspace.cgiar.org/bitstream/handle/10568/69115/CCAFS_INDC_info_note-Final.pdf?sequence=3&isAllowed=y)
34 [Final.pdf?sequence=3&isAllowed=y](https://cgspace.cgiar.org/bitstream/handle/10568/69115/CCAFS_INDC_info_note-Final.pdf?sequence=3&isAllowed=y) (Accessed October 8, 2018).
- 35 Richards, M. B., E. Wollenberg, and D. van Vuuren, 2018: National contributions to climate change
36 mitigation from agriculture: allocating a global target. *Clim. Policy*, **18**, 1271–1285,
37 doi:10.1080/14693062.2018.1430018.
38 <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1430018> (Accessed March 28,
39 2019).
- 40 Richards, P., 2015: What Drives Indirect Land Use Change? How Brazil's Agriculture Sector
41 Influences Frontier Deforestation. *Ann. Assoc. Am. Geogr.*, **105**, 1026–1040,
42 doi:10.1080/00045608.2015.1060924.
43 <http://www.tandfonline.com/doi/full/10.1080/00045608.2015.1060924>.
- 44 Ricroch, A., P. Clairand, and W. Harwood, 2017: Use of CRISPR systems in plant genome editing:
45 toward new opportunities in agriculture. *Emerg. Top. Life Sci.*, **1**, 169–182,
46 doi:10.1042/ETLS20170085.
47 <http://www.emergtoplifesci.org/lookup/doi/10.1042/ETLS20170085> (Accessed April 17, 2019).
- 48 Riley, L., and B. Dodson, 2014: Gendered Mobilities and Food Access in Blantyre, Malawi. *Urban*
49 *Forum*, **25**, 227–239, doi:10.1007/s12132-014-9223-7. [http://link.springer.com/10.1007/s12132-](http://link.springer.com/10.1007/s12132-014-9223-7)

- 1 014-9223-7 (Accessed March 25, 2019).
- 2 Ringler, C., A. R. Quisumbing, E. Bryan, and R. Meinzen-Dick, 2014: *Enhancing Women's Assets to*
3 *Manage Risk Under Climate Change: Potential for Group-based Approaches*. Washington DC,
4 65 pp. <http://ebrary.ifpri.org/cdm/ref/collection/p15738coll2/id/128599> (Accessed November 4,
5 2018).
- 6 Rippke, U., and Coauthors, 2016: Timescales of transformational climate change adaptation in sub-
7 Saharan African agriculture. *Nat. Clim. Chang.*, **6**, 605–609, doi:10.1038/nclimate2947.
8 <http://www.nature.com/doi/10.1038/nclimate2947>.
- 9 Rivera-Ferre, M. G., M. Di Masso, M. Mailhost, F. López-i-Gelats, D. Gallar, I. Vara, and M. Cuellar,
10 2013: Understanding the role of local and traditional agricultural knowledge in a changing world
11 climate: the case of the Indo-Gangetic Plains. *CGIAR-CCAFS Program, Nepal*, **98**.
- 12 Rivera-Ferre, M. G., and Coauthors, 2016a: Local agriculture traditional knowledge to ensure food
13 availability in a changing climate: revisiting water management practices in the Indo-Gangetic
14 Plains. *Agroecol. Sustain. Food Syst.*, **40**, 965–987, doi:10.1080/21683565.2016.1215368.
15 <https://www.tandfonline.com/doi/full/10.1080/21683565.2016.1215368> (Accessed January 11,
16 2018).
- 17 —, F. López-i-Gelats, M. Howden, P. Smith, J. F. Morton, and M. Herrero, 2016b: Re-framing the
18 climate change debate in the livestock sector: mitigation and adaptation options. *Wiley*
19 *Interdiscip. Rev. Clim. Chang.*, **7**, 869–892, doi:10.1002/wcc.421.
- 20 Robinson, L. W., P. J. Ericksen, S. Chesterman, and J. S. Worden, 2015: Sustainable intensification in
21 drylands: What resilience and vulnerability can tell us. *Agric. Syst.*, **135**, 133–140,
22 doi:10.1016/J.AGSY.2015.01.005.
23 <https://www.sciencedirect.com/science/article/pii/S0308521X15000062> (Accessed May 25,
24 2018).
- 25 Robinson, S., and Coauthors, 2014: Comparing supply-side specifications in models of global
26 agriculture and the food system. *Agric. Econ.*, **45**, 21–35, doi:10.1111/agec.12087.
- 27 Robinson, W. I., 2011: Globalization and the sociology of Immanuel Wallerstein: A critical appraisal.
28 *Int. Sociol.*, **26**, 723–745, doi:10.1177/0268580910393372.
29 <http://journals.sagepub.com/doi/10.1177/0268580910393372> (Accessed December 2, 2017).
- 30 Rochedo, P. R. R., and Coauthors, 2018: The threat of political bargaining to climate mitigation in
31 Brazil. *Nat. Clim. Chang.*, **8**, 695–698, doi:10.1038/s41558-018-0213-y.
32 <https://doi.org/10.1038/s41558-018-0213-y>.
- 33 Rockström, J., and Coauthors, 2009: Planetary boundaries: Exploring the safe operating space for
34 humanity. *Ecol. Soc.*, doi:10.5751/ES-03180-140232.
- 35 Rodríguez Osuna, V., J. Börner, and M. Cunha, 2014: Scoping Adaptation Needs for Smallholders in
36 the Brazilian Amazon: A Municipal Level Case Study. *Chang. Adapt. Socio-Ecological Syst.*, **1**,
37 doi:10.2478/cass-2014-0002. [http://www.degruyter.com/view/j/cass.2014.1.issue-1/cass-2014-
38 0002/cass-2014-0002.xml](http://www.degruyter.com/view/j/cass.2014.1.issue-1/cass-2014-0002/cass-2014-0002.xml) (Accessed October 30, 2018).
- 39 Rogelj, J., and Coauthors, 2011: Emission pathways consistent with a 2°C global temperature limit.
40 *Nat. Clim. Chang.*, doi:10.1038/nclimate1258.
- 41 —, G. Luderer, R. C. Pietzcker, E. Kriegler, M. Schaeffer, V. Krey, and K. Riahi, 2015: Energy
42 system transformations for limiting end-of-century warming to below 1.5 °C. *Nat. Clim. Chang.*,
43 **5**, 519–527, doi:10.1038/nclimate2572. <http://dx.doi.org/10.1038/nclimate2572>.
- 44 —, and Coauthors, 2018: Scenarios towards limiting global mean temperature increase below 1.5
45 °C. *Nat. Clim. Chang.*, **8**, 325–332, doi:10.1038/s41558-018-0091-3.
- 46 Rojas-Downing, M. M., A. P. Nejadhashemi, T. Harrigan, and S. A. Woznicki, 2017: Climate change
47 and livestock: Impacts, adaptation, and mitigation. *Clim. Risk Manag.*, **16**, 145–163,
48 doi:10.1016/J.CRM.2017.02.001.

- 1 <https://www.sciencedirect.com/science/article/pii/S221209631730027X> (Accessed March 5,
2 2019).
- 3 Rööös, E., B. Bajželj, P. Smith, M. Patel, D. Little, and T. Garnett, 2017: Greedy or needy? Land use
4 and climate impacts of food in 2050 under different livestock futures. *Glob. Environ. Chang.*,
5 **47**, 1–12, doi:10.1016/j.gloenvcha.2017.09.001.
6 <http://linkinghub.elsevier.com/retrieve/pii/S0959378016306872>.
- 7 Rose, D., M. C. Heller, A. M. Willits-Smith, and R. J. Meyer, 2019: Carbon footprint of self-selected
8 US diets: nutritional, demographic, and behavioral correlates. *Am. J. Clin. Nutr.*, **109**, 526–534,
9 doi:10.1093/ajcn/nqy327.
- 10 Rose, T. J., R. H. Wood, M. T. Rose, and L. Van Zwieten, 2018: A re-evaluation of the agronomic
11 effectiveness of the nitrification inhibitors DCD and DMPP and the urease inhibitor NBPT.
12 *Agric. Ecosyst. Environ.*, **252**, 69–73, doi:10.1016/J.AGEE.2017.10.008.
13 <https://www.sciencedirect.com/science/article/pii/S0167880917304541> (Accessed May 25,
14 2018).
- 15 Rosegrant, M. W., E. Magalhaes, R. A. Valmonte-Santos, and D. Mason-D’Croz, 2015: *Returns to*
16 *investment in reducing postharvest food losses and increasing agricultural productivity growth:*
17 *Post-2015 consensus*. Lowell, MA, USA: Copenhagen Consensus Center, 46 pp.
18 [http://www.copenhagenconsensus.com/sites/default/files/food_security_nutrition_assessment_-](http://www.copenhagenconsensus.com/sites/default/files/food_security_nutrition_assessment_-_rosegrant_0.pdf)
19 [_rosegrant_0.pdf](http://www.copenhagenconsensus.com/sites/default/files/food_security_nutrition_assessment_-_rosegrant_0.pdf) (Accessed September 26, 2018).
- 20 Rosemberg, A., 2010: Building a Just Transition: The linkages between climate change and
21 employment. *Int. J. Labour Res.*, **2**, 125–161.
22 <https://search.proquest.com/docview/884976739?pq-origsite=gscholar> (Accessed March 25,
23 2019).
- 24 Rosenstock, T., K. Tully, C. Arias-Navarro, H. Neufeldt, K. Butterbach-Bahl, and L. Verchot, 2014:
25 Agroforestry with N₂-fixing trees: sustainable development’s friend or foe? *Curr. Opin.*
26 *Environ. Sustain.*, **6**, 15–21, doi:10.1016/j.cosust.2013.09.001.
27 <http://linkinghub.elsevier.com/retrieve/pii/S1877343513001218> (Accessed May 25, 2018).
- 28 Rosenzweig, C., and D. Hillel, 2015: *Handbook of Climate Change and Agroecosystems: The*
29 *Agricultural Model Intercomparison and Improvement Project (AgMIP) Integrated Crop and*
30 *Economic Assessments*. First. World Scientific,.
- 31 —, and W. Solecki, 2018: Action pathways for transforming cities. *Nat. Clim. Chang.*, **8**, 756–759,
32 doi:10.1038/s41558-018-0267-x. <http://www.nature.com/articles/s41558-018-0267-x> (Accessed
33 October 1, 2018).
- 34 —, and Coauthors, 2014: Assessing agricultural risks of climate change in the 21st century in a
35 global gridded crop model intercomparison. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 3268–3273,
36 doi:10.1073/pnas.1222463110.
37 <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=3948251&tool=pmcentrez&renderty>
38 [pe=abstract](http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=3948251&tool=pmcentrez&renderty).
- 39 —, and Coauthors, 2018a: Coordinating AgMIP data and models across global and regional scales
40 for 1.5 °C and 2.0 °C assessments. *Phil. Trans. R. Soc.*, doi:10.1098/rsta.2016.0455.
- 41 Rosenzweig, C., W. Solecki, P. Romero-Lankao, S. Mehrotra, S. Dhakal, T. Bowman, and S. A.
42 Ibrahim, 2018b: *Climate Change and Cities: Second Assessment Report of the Urban Climate*
43 *Change Research Network*.
- 44 Rosi, A., and Coauthors, 2017: Environmental impact of omnivorous, ovo-lacto-vegetarian, and vegan
45 diet. *Sci. Rep.*, **7**, 1–9, doi:10.1038/s41598-017-06466-8. [http://dx.doi.org/10.1038/s41598-017-](http://dx.doi.org/10.1038/s41598-017-06466-8)
46 [06466-8](http://dx.doi.org/10.1038/s41598-017-06466-8).
- 47 Ross, C. L., M. Orenstein, and N. Botchwey, 2014: Public Health and Community Planning 101.
48 *Health Impact Assessment in the United States*, Springer New York, New York, NY, 15–32
49 http://link.springer.com/10.1007/978-1-4614-7303-9_2 (Accessed November 5, 2018).

- 1 Rossi, A., and Y. Lambrou, 2008: *Gender and equity issues in liquid biofuels production: Minimizing*
2 *the risks to maximize the opportunities*. FAO, Rome, 33 pp.
3 <http://www.fao.org/docrep/pdf/010/i0143e/i0143e00.pdf> (Accessed May 31, 2018).
- 4 Roy, J., and Coauthors, 2018: Sustainable Development , Poverty Eradication and Reducing
5 Inequalities. *Global Warming of 1.5 °C an IPCC special report on the impacts of global*
6 *warming of 1.5 °C above pre-industrial levels and related global greenhouse gas emission*
7 *pathways, in the context of strengthening the global response to the threat of climate change*
8 <http://www.ipcc.ch/report/sr15/>.
- 9 Ruby, M. B., and S. J. Heine, 2011: Meat, morals, and masculinity. *Appetite*, **56**, 447–450,
10 doi:10.1016/J.APPET.2011.01.018.
11 <https://www.sciencedirect.com/science/article/pii/S0195666311000341> (Accessed April 5,
12 2019).
- 13 Ruel-Bergeron, J. C., G. A. Stevens, J. D. Sugimoto, F. F. Roos, M. Ezzati, R. E. Black, and K.
14 Kraemer, 2015: Global update and trends of hidden hunger, 1995-2011: The hidden hunger
15 Index. *PLoS One*, **10**, 1–13, doi:10.1371/journal.pone.0143497.
16 <http://dx.doi.org/10.1371/journal.pone.0143497>.
- 17 Ruel, M. T., J. L. Garrett, C. Hawkes, and M. J. Cohen, 2010: The Food, Fuel, and Financial Crises
18 Affect the Urban and Rural Poor Disproportionately: A Review of the Evidence. *J. Nutr.*, **140**,
19 170S–176S, doi:10.3945/jn.109.110791.
20 <https://academic.oup.com/jn/article/140/1/170S/4743305> (Accessed May 25, 2018).
- 21 Ruiter, M. De, P. Hudson, L. De Ruig, O. Kuik, and W. Botzen, 2017: A comparative study of
22 European insurance schemes for extreme weather risks and incentives for risk reduction. *EGU*
23 *General Assembly*, Vol. 19 of, EGU.
- 24 Rumpold, B. A., and O. Schlüter, 2015: Insect-based protein sources and their potential for human
25 consumption: Nutritional composition and processing. *Anim. Front.*, **5**, 20–24,
26 doi:10.2527/af.2015-0015. <https://academic.oup.com/af/article/5/2/20/4638722> (Accessed April
27 23, 2019).
- 28 Ryan, M. R., T. E. Crews, S. W. Culman, L. R. Dehaan, R. C. Hayes, J. M. Jungers, and M. G.
29 Bakker, 2018: Managing for Multifunctionality in Perennial Grain Crops. *Bioscience*,
30 doi:10.1093/biosci/biy014.
- 31 Saberi, S., M. Kouhizadeh, J. Sarkis, and L. Shen, 2019: Blockchain technology and its relationships
32 to sustainable supply chain management. *Int. J. Prod. Res.*, **57**, 2117–2135,
33 doi:10.1080/00207543.2018.1533261.
34 <https://www.tandfonline.com/doi/full/10.1080/00207543.2018.1533261> (Accessed April 17,
35 2019).
- 36 Saki, M., M. Tarkesh Esfahani, and S. Soltani, 2018: A scenario-based modeling of climate change
37 impacts on the aboveground net primary production in rangelands of central Iran. *Environ. Earth*
38 *Sci.*, **77**, 670, doi:10.1007/s12665-018-7864-x. [http://link.springer.com/10.1007/s12665-018-](http://link.springer.com/10.1007/s12665-018-7864-x)
39 [7864-x](http://link.springer.com/10.1007/s12665-018-7864-x) (Accessed April 25, 2019).
- 40 Salou, T., C. Le Mouël, and H. M. G. van der Werf, 2017: Environmental impacts of dairy system
41 intensification: the functional unit matters! *J. Clean. Prod.*, **140**, 445–454,
42 doi:10.1016/J.JCLEPRO.2016.05.019.
43 <https://www.sciencedirect.com/science/article/pii/S0959652616304747> (Accessed April 5,
44 2019).
- 45 Saltiel, A. R., and J. M. Olefsky, 2017: Inflammatory mechanisms linking obesity and metabolic
46 disease. *J. Clin. Invest.*, **127**, 1–4, doi:10.1172/JCI92035.
47 <https://www.jci.org/articles/view/92035> (Accessed December 2, 2017).
- 48 Salvador, S., M. Corazzin, A. Romanzin, and S. Bovolenta, 2017: Greenhouse gas balance of
49 mountain dairy farms as affected by grassland carbon sequestration. *J. Environ. Manage.*, **196**,

- 1 644–650, doi:10.1016/J.JENVMAN.2017.03.052.
2 <https://www.sciencedirect.com/science/article/pii/S0301479717302670> (Accessed April 16,
3 2019).
- 4 Samy, A. M., and A. T. Peterson, 2016: Climate Change Influences on the Global Potential
5 Distribution of Bluetongue Virus. *PLoS One*, **11**, e0150489, doi:10.1371/journal.pone.0150489.
6 <https://doi.org/10.1371/journal.pone.0150489>.
- 7 Sanderson, M. A., G. Brink, R. Stout, and L. Ruth, 2013: Grass–Legume Proportions in Forage Seed
8 Mixtures and Effects on Herbage Yield and Weed Abundance. *Agron. J.*, **105**, 1289,
9 doi:10.2134/agronj2013.0131.
- 10 Sans, P., and P. Combris, 2015: World meat consumption patterns: An overview of the last fifty years
11 (1961–2011). *Meat Sci.*, **109**, 106–111, doi:10.1016/j.meatsci.2015.05.012.
- 12 Sanz, M. J., and Coauthors, 2017: *Sustainable Land Management contribution to successful land-*
13 *based climate change adaptation and mitigation. A Report of the Science-Policy Interface.*
14 Bonn, Germany, 178 pp. [https://www.unccd.int/sites/default/files/documents/2017-](https://www.unccd.int/sites/default/files/documents/2017-09/UNCCD_Report_SLM_web_v2.pdf)
15 [09/UNCCD_Report_SLM_web_v2.pdf](https://www.unccd.int/sites/default/files/documents/2017-09/UNCCD_Report_SLM_web_v2.pdf).
- 16 Sapkota, T., and Coauthors, 2017a: Soil organic carbon changes after seven years of conservation
17 agriculture based rice-wheat cropping system in the eastern Indo-Gangetic Plain of India. *Soil*
18 *Use Manag.*, 1–9, doi:10.1111/sum.12331.
- 19 Sapkota, T. B., M. L. Jat, J. P. Aryal, R. K. Jat, and A. Khatri-Chhetri, 2015: Climate change
20 adaptation, greenhouse gas mitigation and economic profitability of conservation agriculture:
21 Some examples from cereal systems of Indo-Gangetic Plains. *J. Integr. Agric.*, **14**, 1524–1533,
22 doi:10.1016/S2095-3119(15)61093-0.
- 23 Sapkota, T. B., J. P. Aryal, A. Khatri-chhetri, P. B. Shirsath, P. Arumugam, and C. M. Stirling, 2017b:
24 Identifying high-yield low-emission pathways for the cereal production in South Asia. *Mitig.*
25 *Adapt. Strateg. Glob. Chang.*, doi:10.1007/s11027-017-9752-1.
- 26 Sapkota, T. B., V. Shankar, M. Rai, M. L. Jat, C. M. Stirling, L. K. Singh, H. S. Jat, and M. S.
27 Grewal, 2017c: Reducing global warming potential through sustainable intensification of
28 Basmati rice-wheat systems in India. *Sustain.*, **9**, doi:10.3390/su9061044.
- 29 Sapkota, T. B. T. B., K. Majumdar, M. L. Jat, A. Kumar, D. K. Bishnoi, A. J. McDonald, and M.
30 Pampolino, 2014: Precision nutrient management in conservation agriculture based wheat
31 production of Northwest India: Profitability, nutrient use efficiency and environmental footprint.
32 *F. Crop. Res.*, **155**, 233–244, doi:10.1016/j.fcr.2013.09.001.
- 33 Saxena, A. K., X. C. Fuentes, R. G. Herbas, and D. L. Humphries, 2016: indigenous Food systems
34 and climate change: impacts of climatic shifts on the Production and Processing of native and
35 Traditional crops in the Bolivian andes. *Front. public Heal.*, **4**.
- 36 Sayer, J., and Coauthors, 2013: Ten principles for a landscape approach to reconciling agriculture,
37 conservation, and other competing land uses. *Proc. Natl. Acad. Sci.*,
38 doi:10.1073/pnas.1210595110.
- 39 Scanlon, B. R., B. L. Ruddell, P. M. Reed, R. I. Hook, C. Zheng, V. C. Tidwell, and S. Siebert, 2017:
40 The food-energy-water nexus: Transforming science for society. *Water Resour. Res.*, **53**, 3550–
41 3556, doi:10.1002/2017WR020889.
- 42 Scheba, A., 2017: Conservation agriculture and sustainable development in Africa: Insights from
43 Tanzania. *Nat. Resour. Forum*, doi:10.1111/1477-8947.12123.
- 44 Scheelbeek, P. F. D., and Coauthors, 2018: Effect of environmental changes on vegetable and legume
45 yields and nutritional quality. *Proc. Natl. Acad. Sci.*, **115**, 6804 LP-6809.
46 <http://www.pnas.org/content/115/26/6804.abstract>.
- 47 Schenker, S., 2003: Undernutrition in the UK. *Bull. Nutr. / Br. Nutr. Found.*, **28**, 87–120,
48 doi:10.1046/j.1467-3010.2003.00303.x. <http://doi.wiley.com/10.1046/j.1467-3010.2003.00303.x>

- 1 (Accessed December 1, 2017).
- 2 Schleussner, C.-F., J. F. Donges, R. V. Donner, and H. J. Schellnhuber, 2016: Armed-conflict risks
3 enhanced by climate-related disasters in ethnically fractionalized countries. *Proc. Natl. Acad.*
4 *Sci.*, **113**, 9216–9221, doi:10.1073/PNAS.1601611113.
5 <http://www.pnas.org/content/113/33/9216> (Accessed November 10, 2018).
- 6 ———, and Coauthors, 2018: Crop productivity changes in 1.5 °C and 2 °C worlds under climate
7 sensitivity uncertainty. *Environ. Res. Lett.*, **13**, 64007, doi:10.1088/1748-9326/aab63b.
8 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=6/a=064007?key=crossref.ee5fe2d10bbd8a6808c650861a04f82e)
9 [9326/13/i=6/a=064007?key=crossref.ee5fe2d10bbd8a6808c650861a04f82e](http://stacks.iop.org/1748-9326/13/i=6/a=064007?key=crossref.ee5fe2d10bbd8a6808c650861a04f82e).
- 10 Schmidhuber, J., and F. N. Tubiello, 2007: Global food security under climate change. *Proc. Natl.*
11 *Acad. Sci. U. S. A.*, **104**, 19703–19708, doi:10.1073/pnas.0701976104.
12 <http://www.pnas.org/content/104/50/19703.short>.
- 13 Schmidtko, S., L. Stramma, and M. Visbeck, 2017: Decline in global oceanic oxygen content during
14 the past five decades. *Nature*, **542**, 335.
- 15 Schmitz, C., and Coauthors, 2014: Land-use change trajectories up to 2050: insights from a global
16 agro-economic model comparison. *Agric. Econ.*, **45**, 69–84, doi:10.1111/agec.12090.
- 17 Schmitz, O. J., and Coauthors, 2015: Conserving Biodiversity : Practical Guidance about Climate
18 Change Adaptation Approaches in Support of Land- use Planning. *Nat. Areas J.*, **1**, 190–203.
- 19 Schramski, S., C. Mccarty, and G. Barnes, 2017: Household adaptive capacity: a social networks
20 approach in rural South Africa. *Clim. Dev.*, **0**, 1–13, doi:10.1080/17565529.2017.1301861.
- 21 Schweiger, O., and Coauthors, 2010: Multiple stressors on biotic interactions: how climate change
22 and alien species interact to affect pollination. *Biol. Rev.*, **85**, 777–795, doi:10.1111/j.1469-
23 185X.2010.00125.x.
- 24 Scott, P., B. Tayler, and D. Walters, 2017: Lessons from implementing integrated water resource
25 management: a case study of the North Bay-Mattawa Conservation Authority, Ontario. *Int. J.*
26 *Water Resour. Dev.*, **33**, 393–407, doi:10.1080/07900627.2016.1216830.
27 <http://dx.doi.org/10.1080/07900627.2016.1216830>.
- 28 Searchinger, T. D., S. Wirsenius, T. Beringer, and P. Dumas, 2018: Assessing the efficiency of
29 changes in land use for mitigating climate change. *Nature*, **564**, 249–253, doi:10.1038/s41586-
30 018-0757-z.
- 31 Sehmi, R., C. Mbow, S. Pitkanen, H. Cross, N. Berry, M. Riddell, J. Heiskanen, and E. Aynekulu,
32 2016: Replicable tools and frameworks for biocarbon development in West Africa. *ICRAF*
33 *Work. Pap. - World Agrofor. Cent.*, xi + 74 pp.
- 34 Selby, J., O. S. Dahi, C. Fröhlich, and M. Hulme, 2017: Climate change and the Syrian civil war
35 revisited. *Polit. Geogr.*, **60**, 232–244, doi:10.1016/J.POLGEO.2017.05.007.
36 <https://www.sciencedirect.com/science/article/pii/S0962629816301822> (Accessed May 22,
37 2018).
- 38 Senay, G. B., N. M. Velpuri, S. Bohms, M. Budde, C. Young, J. Rowland, and J. P. Verdin, 2015:
39 Drought Monitoring and Assessment : Remote Sensing and Modeling Approaches for the
40 Famine Early Warning Systems Network. *Hydro-Meteorological Hazards, Risks and Disasters*,
41 Elsevier, 233–262 <http://linkinghub.elsevier.com/retrieve/pii/B9780123948465000096>.
- 42 Sepulcre-Canto, G., J. Vogt, A. Arboleda, and T. Antofie, 2014: Assessment of the EUMETSAT
43 LSA-SAF evapotranspiration product for drought monitoring in Europe. *Int. J. Appl. Earth Obs.*
44 *Geoinf.*, **30**, 190–202, doi:10.1016/j.jag.2014.01.021.
45 <http://dx.doi.org/10.1016/j.jag.2014.01.021>.
- 46 Settele, J., J. Bishop, and S. G. Potts, 2016: Climate change impacts on pollination. *Nat. Plants*, **2**,
47 doi:10.1038/nplants.2016.92.

- 1 Shah, M., G. Fischer, and H. Van Velthuizen, 2008: *Food security and sustainable agriculture: The*
2 *challenges of climate change in Sub-Saharan Africa*. Laxenburg, 40 pp.
- 3 Shepon, A., G. Eshel, E. Noor, and R. Milo, 2016: Energy and protein feed-to-food conversion
4 efficiencies in the US and potential food security gains from dietary changes. *Environ. Res.*
5 *Lett.*, **11**, doi:10.1088/1748-9326/11/10/105002.
- 6 Shepon, A., G. Eshel, E. Noor, and R. Milo, 2018: The opportunity cost of animal based diets exceeds
7 all food losses. *Proc. Natl. Acad. Sci.*, **115**, 3804 LP-3809.
8 <http://www.pnas.org/content/115/15/3804.abstract>.
- 9 Shi, Q., Y. Lin, E. Zhang, H. Yan, and J. Zhan, 2013: Impacts of Cultivated Land Reclamation on the
10 Climate and Grain Production in Northeast China in the Future 30 Years. *Adv. in Meteorology*,
11 **2013**.
- 12 Shindell, D., and Coauthors, 2012: Simultaneously mitigating near-term climate change and
13 improving human health and food security. *Science*, **335**, 183–189,
14 doi:10.1126/science.1210026. <http://www.ncbi.nlm.nih.gov/pubmed/22246768> (Accessed April
15 16, 2019).
- 16 Shindell, D. T., 2016: Crop yield changes induced by emissions of individual climate-altering
17 pollutants. *Earth's Futur.*, **4**, 373–380.
- 18 Shindell, D. T., J. S. Fuglestedt, and W. J. Collins, 2017: The social cost of methane: theory and
19 applications. *Faraday Discuss.*, **200**, 429–451, doi:10.1039/C7FD00009J.
20 <http://xlink.rsc.org/?DOI=C7FD00009J>.
- 21 Shove, E., M. Pantzar, and M. Watson, 2012: *The Dynamics of Social Practice: Everyday Life and*
22 *How it Changes*. SAGE Publications Ltd, London, UK, [http://sk.sagepub.com/books/the-](http://sk.sagepub.com/books/the-dynamics-of-social-practice)
23 [dynamics-of-social-practice](http://sk.sagepub.com/books/the-dynamics-of-social-practice) (Accessed November 5, 2018).
- 24 Shrestha, R. P., and N. Nepal, 2016: An assessment by subsistence farmers of the risks to food
25 security attributable to climate change in Makwanpur, Nepal. *Food Secur.*, **8**, 415–425.
- 26 Sibhatu, K. T., V. V. Krishna, and M. Qaim, 2015: Production diversity and dietary diversity in
27 smallholder farm households. *Proc. Natl. Acad. Sci.*, **112**, 10657–10662,
28 doi:10.1073/pnas.1510982112. <http://www.pnas.org/lookup/doi/10.1073/pnas.1510982112>.
- 29 Sida, T. S., F. Baudron, H. Kim, and K. E. Giller, 2018: Climate-smart agroforestry: Faidherbia albida
30 trees buffer wheat against climatic extremes in the Central Rift Valley of Ethiopia. *Agric. For.*
31 *Meteorol.*, **248**, 339–347, doi:10.1016/J.AGRFORMET.2017.10.013.
32 <https://www.sciencedirect.com/science/article/pii/S0168192317303337> (Accessed March 9,
33 2019).
- 34 Sietz, D., S. E. Mamani Choque, and M. K. B. Lüdeke, 2012: Typical patterns of smallholder
35 vulnerability to weather extremes with regard to food security in the Peruvian Altiplano. *Reg.*
36 *Environ. Chang.*, **12**, 489–505, doi:10.1007/s10113-011-0246-5.
37 <http://link.springer.com/10.1007/s10113-011-0246-5> (Accessed March 30, 2019).
- 38 Siliprandi, E., and G. P. Zuluaga Sánchez, 2014: *Género, agroecología y soberanía alimentaria :*
39 *perspectivas ecofeministas*. Icaria, <http://www.icariaeditorial.com/libros.php?id=1475> (Accessed
40 April 13, 2019).
- 41 Silva Oliveira, R., L. G. Barioni, J. A. J. Hall, A. C. Moretti, R. Fonseca Veloso, P. Alexander, M.
42 Crespolini, and D. Moran, 2017: Sustainable intensification of Brazilian livestock production
43 through optimized pasture restoration. *Agric. Syst.*, **153**, 201–211,
44 doi:10.1016/J.AGSY.2017.02.001.
45 <https://www.sciencedirect.com/science/article/pii/S0308521X16303845> (Accessed June 2,
46 2018).
- 47 Singh, P., K. J. Boote, M. D. M. Kadiyala, S. Nedumaran, S. K. Gupta, K. Srinivas, and M. C. S.
48 Bantilan, 2017: An assessment of yield gains under climate change due to genetic modification

- 1 of pearl millet. *Sci. Total Environ.*, **601–602**, 1226–1237, doi:10.1016/j.scitotenv.2017.06.002.
- 2 Skakun, S., N. Kussul, A. Shelestov, and O. Kussul, 2017: The use of satellite data for agriculture
3 drought risk quantification in Ukraine. **5705**, doi:10.1080/19475705.2015.1016555.
- 4 Skelsey, P., and A. C. Newton, 2015: Future environmental and geographic risks of Fusarium head
5 blight of wheat in Scotland. *Eur. J. Plant Pathol.*, **142**, 133–147, doi:10.1007/s10658-015-0598-
6 7.
- 7 Skoufias, E., M. Rabassa, and S. Olivieri, 2011: *The poverty impacts of climate change: A review of*
8 *the evidence*. The World Bank, <http://elibrary.worldbank.org/doi/book/10.1596/1813-9450-5622>
9 (Accessed March 20, 2019).
- 10 Slade, P., 2018: If you build it, will they eat it? Consumer preferences for plant-based and cultured
11 meat burgers. *Appetite*, **125**, 428–437, doi:10.1016/J.APPET.2018.02.030.
12 <https://www.sciencedirect.com/science/article/pii/S0195666317317531> (Accessed November 4,
13 2018).
- 14 Smetana, S., A. Mathys, A. Knoch, and V. Heinz, 2015: Meat alternatives: life cycle assessment of
15 most known meat substitutes. *Int. J. Life Cycle Assess.*, **20**, 1254–1267, doi:10.1007/s11367-
16 015-0931-6.
- 17 Smith, P., 2013: Delivering food security without increasing pressure on land. *Glob. Food Sec.*, **2**, 18–
18 23, doi:10.1016/j.gfs.2012.11.008. <http://dx.doi.org/10.1016/j.gfs.2012.11.008>.
- 19 —, and Coauthors, 2013: How much land-based greenhouse gas mitigation can be achieved without
20 compromising food security and environmental goals? *Glob. Chang. Biol.*, **19**, 2285–2302,
21 doi:10.1111/gcb.12160.
- 22 Smith, P., and Coauthors, 2014: Agriculture, Forestry and Other Land Use (AFOLU). *Climate*
23 *Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth*
24 *Assessment Report of the Intergovernmental Panel on Climate Change*, O. Edenhofer et al.,
25 Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA
26 http://www.ipcc-nggip.iges.or.jp/public/2006gl/pdf/4_Volume4/V4_04_Ch4_Forest_Land.pdf.
- 27 Smith, P., and Coauthors, 2016: Global change pressures on soils from land use and management.
28 *Glob. Chang. Biol.*, **22**, 1008–1028, doi:10.1111/gcb.13068.
29 <http://doi.wiley.com/10.1111/gcb.13068> (Accessed October 1, 2018).
- 30 Smithers, G. W., 2015: Food Science – Yesterday, Today and Tomorrow. *Reference Module in Food*
31 *Science*, Elsevier <http://linkinghub.elsevier.com/retrieve/pii/B9780081005965033370>.
- 32 Soares-Filho, B., R. Rajão, M. Macedo, A. Carneiro, W. Costa, M. Coe, H. Rodrigues, and A.
33 Alencar, 2014: Land use. Cracking Brazil’s Forest Code. *Science*, **344**, 363–364,
34 doi:10.1126/science.1246663. <http://www.ncbi.nlm.nih.gov/pubmed/24763575> (Accessed May
35 31, 2018).
- 36 Sommer, M., S. Ferron, S. Cavill, and S. House, 2015: Violence, gender and WASH: spurring action
37 on a complex, under-documented and sensitive topic. *Environ. Urban.*, **27**, 105–116,
38 doi:10.1177/0956247814564528. <http://journals.sagepub.com/doi/10.1177/0956247814564528>
39 (Accessed April 6, 2019).
- 40 Song, G., M. Li, P. Fullana-i-Palmer, D. Williamson, and Y. Wang, 2017: Dietary changes to mitigate
41 climate change and benefit public health in China. *Sci. Total Environ.*, **577**, 289–298,
42 doi:10.1016/j.scitotenv.2016.10.184. <http://dx.doi.org/10.1016/j.scitotenv.2016.10.184>.
- 43 Sousa, V. M. Z., S. M. Luz, A. Caldeira-Pires, F. S. Machado, and C. M. Silveira, 2017: Life cycle
44 assessment of biodiesel production from beef tallow in Brazil. *Int. J. Life Cycle Assess.*, **22**,
45 1837–1850, doi:10.1007/s11367-017-1396-6.
- 46 Soussana, J., L. G. Barioni, T. Ben Ari, R. Conant, P. Gerber, P. Havlík, A. Ickowicz, and M. H. G.,
47 2013: Managing grassland systems in a changing climate: the search for practical solutions.
48 *Revitalising grasslands to sustain our communities, 22nd International Grasslands Congress, 15*

- 1 - 19 September 2013, Sidney (Australia), 1–18.
- 2 Soussana, J.-F., and Coauthors, 2019: Matching policy and science: Rationale for the “4 per 1000 -
3 soils for food security and climate” initiative. *Soil Tillage Res.*, **188**, 3–15,
4 doi:10.1016/J.STILL.2017.12.002.
- 5 de Souza, N. R. D., and Coauthors, 2019: Sugarcane ethanol and beef cattle integration in Brazil.
6 *Biomass and Bioenergy*, **120**, 448–457, doi:10.1016/J.BIOMBIOE.2018.12.012.
- 7 Spence, L. A., P. Liancourt, B. Boldgiv, P. S. Petraitis, and B. B. Casper, 2014: Climate change and
8 grazing interact to alter flowering patterns in the Mongolian steppe. *Oecologia*, **175**, 251–260,
9 doi:10.1007/s00442-014-2884-z. <http://link.springer.com/10.1007/s00442-014-2884-z>
10 (Accessed March 15, 2019).
- 11 van der Spiegel, M., H. J. van der Fels-Klerx, and H. J. P. Marvin, 2012: Effects of climate change on
12 food safety hazards in the dairy production chain. *Food Res. Int.*, **46**, 201–208,
13 doi:10.1016/j.foodres.2011.12.011.
- 14 Spires, M., S. Shackleton, and G. Cundill, 2014: Barriers to implementing planned community-based
15 adaptation in developing countries: a systematic literature review. *Clim. Dev.*, **6**, 277–287,
16 doi:10.1080/17565529.2014.886995.
17 <http://www.tandfonline.com/doi/abs/10.1080/17565529.2014.886995> (Accessed October 24,
18 2018).
- 19 Spratt, S., 2013: *Food price volatility and financial speculation*. 22 pp.
20 <https://pdfs.semanticscholar.org/3af2/d8cf9fdc5a15131e1002757a1428d950de49.pdf>.
- 21 Springmann, M., and Coauthors, 2016a: Global and regional health effects of future food production
22 under climate change: A modelling study. *Lancet*, **387**, 1937–1946, doi:10.1016/S0140-
23 6736(15)01156-3. [http://dx.doi.org/10.1016/S0140-6736\(15\)01156-3](http://dx.doi.org/10.1016/S0140-6736(15)01156-3).
- 24 —, H. C. J. Godfray, M. Rayner, and P. Scarborough, 2016b: Analysis and valuation of the health
25 and climate change cobenefits of dietary change. *Proc. Natl. Acad. Sci.*, **113**, 4146–4151,
26 doi:10.1073/pnas.1523119113. <http://www.pnas.org/lookup/doi/10.1073/pnas.1523119113>.
- 27 —, D. Mason-D’Croz, S. Robinson, K. Wiebe, H. C. J. Godfray, M. Rayner, and P. Scarborough,
28 2017: Mitigation potential and global health impacts from emissions pricing of food
29 commodities. *Nat. Clim. Chang.*, **7**, 69–74, doi:10.1038/nclimate3155.
30 <http://www.nature.com/doi/10.1038/nclimate3155>.
- 31 —, and Coauthors, 2018a: Options for keeping the food system within environmental limits.
32 *Nature*, doi:10.1038/s41586-018-0594-0. <https://doi.org/10.1038/s41586-018-0594-0>.
- 33 —, K. Wiebe, D. Mason-D’Croz, T. B. Sulser, M. Rayner, and P. Scarborough, 2018b: Health and
34 nutritional aspects of sustainable diet strategies and their association with environmental
35 impacts: a global modelling analysis with country-level detail. *Lancet. Planet. Heal.*, **2**, e451–
36 e461, doi:10.1016/S2542-5196(18)30206-7. <http://www.ncbi.nlm.nih.gov/pubmed/30318102>
37 (Accessed November 5, 2018).
- 38 Srbínovska, M., C. Gavrovski, V. Dimcev, A. Krkoleva, and V. Borozan, 2015: Environmental
39 parameters monitoring in precision agriculture using wireless sensor networks. *J. Clean. Prod.*,
40 doi:10.1016/j.jclepro.2014.04.036.
- 41 Srinivasa Rao, C., K. A. Gopinath, J. V. N. S. Prasad, Prasannakumar, and A. K. Singh, 2016:
42 Climate Resilient Villages for Sustainable Food Security in Tropical India: Concept, Process,
43 Technologies, Institutions, and Impacts. *Adv. Agron.*, **140**, 101–214,
44 doi:10.1016/BS.AGRON.2016.06.003.
45 <https://www.sciencedirect.com/science/article/pii/S0065211316300797> (Accessed April 17,
46 2019).
- 47 Stehfest, E., L. Bouwman, D. P. Van Vuuren, M. G. J. Den Elzen, B. Eickhout, and P. Kabat, 2009:
48 Climate benefits of changing diet. *Clim. Change*, **95**, 83–102, doi:10.1007/s10584-008-9534-6.

- 1 ———, and Coauthors, 2019: Key determinants of global land-use futures. *Nat. Commun.*, **Accepted**.
- 2 Steinfeld, H., 2010: *Livestock in a changing landscape. Volume 1, Drivers, consequences, and*
3 *responses*. Island Press, 396 pp.
- 4 Stephens, N., L. Di Silvio, I. Dunsford, M. Ellis, A. Glencross, and A. Sexton, 2018: Bringing
5 cultured meat to market: Technical, socio-political, and regulatory challenges in cellular
6 agriculture. *Trends Food Sci. Technol.*, **78**, 155–166, doi:10.1016/J.TIFS.2018.04.010.
7 <https://www.sciencedirect.com/science/article/pii/S0924224417303400> (Accessed November 4,
8 2018).
- 9 Stojanov, R., B. Duží, I. Kelman, D. Němec, and D. Procházka, 2017: Local perceptions of climate
10 change impacts and migration patterns in Malé, Maldives. *Geogr. J.*, **183**, 370–385,
11 doi:10.1111/geoj.12177. <http://doi.wiley.com/10.1111/geoj.12177> (Accessed May 8, 2018).
- 12 Strassburg, B. B. N., A. E. Latawiec, L. G. Barioni, C. a. Nobre, V. P. da Silva, J. F. Valentim, M.
13 Vianna, and E. D. Assad, 2014: When enough should be enough: Improving the use of current
14 agricultural lands could meet production demands and spare natural habitats in Brazil. *Glob.*
15 *Environ. Chang.*, **28**, 84–97, doi:10.1016/j.gloenvcha.2014.06.001.
- 16 Subash, N., H. Singh, B. Gangwar, G. Baigorria, A. K. Sikka, and R. O. Valdivia, 2015: Integrated
17 Climate Change Assessment through Linking Crop Simulation with Economic Modeling—
18 Results from the Indo-Gangetic Basin. *Handbook of Climate Change and agroecosystems: The*
19 *Agricultural Model Intercomparison and Improvement Project Integrated Crop and Economic*
20 *Assessments, Part 2*, World Scientific, 259–280.
- 21 Subbarao, G. V, and Coauthors, 2009: Evidence for biological nitrification inhibition in Brachiaria
22 pastures. *Proc. Natl. Acad. Sci.*, **106**, 17302–17307.
- 23 Sugiura, T., H. Ogawa, N. Fukuda, and T. Moriguchi, 2013: Changes in the taste and textural
24 attributes of apples in response to climate change. *Sci. Rep.*, **3**, 2418, doi:10.1038/srep02418.
25 <http://www.nature.com/articles/srep02418> (Accessed April 13, 2019).
- 26 Sultan, B., and M. Gaetani, 2016: Agriculture in West Africa in the Twenty-First Century: Climate
27 Change and Impacts Scenarios, and Potential for Adaptation. *Front. Plant Sci.*, **7**, 1–20,
28 doi:10.3389/fpls.2016.01262.
29 <http://journal.frontiersin.org/Article/10.3389/fpls.2016.01262/abstract>.
- 30 Swinburn, B. A., and Coauthors, 2019: The Global Syndemic of Obesity, Undernutrition, and Climate
31 Change: The Lancet Commission report. *Lancet (London, England)*, **393**, 791–846,
32 doi:10.1016/S0140-6736(18)32822-8. <http://www.ncbi.nlm.nih.gov/pubmed/30700377>
33 (Accessed March 20, 2019).
- 34 Tadasse, G., B. Algieri, M. Kalkuhl, and J. von Braun, 2016: Drivers and Triggers of International
35 Food Price Spikes and Volatility. *Food Price Volatility and Its Implications for Food Security*
36 *and Policy*, Springer International Publishing, Cham, 59–82
37 http://link.springer.com/10.1007/978-3-319-28201-5_3 (Accessed December 2, 2017).
- 38 Takakura, J., S. Fujimori, K. Takahashi, and M. Hijioka, Yasuaki and Tomoko Hasegawa and Yasushi
39 HondaToshihiko, 2017: Cost of preventing workplace heat-related illness through worker breaks
40 and the benefit of climate-change mitigation. *Environ. Res. Lett.*, **12**, 64010.
41 <http://stacks.iop.org/1748-9326/12/i=6/a=064010>.
- 42 Tao, F., and Coauthors, 2014: Responses of wheat growth and yield to climate change in different
43 climate zones of China, 1981–2009. *Agric. For. Meteorol.*, **189–190**, 91–104,
44 doi:10.1016/J.AGRFORMET.2014.01.013.
45 <https://www.sciencedirect.com/science/article/pii/S0168192314000227?via%3Dihub> (Accessed
46 April 12, 2019).
- 47 Tariq, M., and Coauthors, 2018: The impact of climate warming and crop management on phenology
48 of sunflower-based cropping systems in Punjab, Pakistan. *Agric. For. Meteorol.*, **256–257**, 270–
49 282, doi:10.1016/J.AGRFORMET.2018.03.015.

- 1 <https://www.sciencedirect.com/science/article/pii/S0168192318300984> (Accessed November 4,
2 2018).
- 3 Taub, D. R., and X. Wang, 2008: Why are nitrogen concentrations in plant tissues lower under
4 elevated CO₂? A critical examination of the hypotheses. *J. Integr. Plant Biol.*, **50**, 1365–1374,
5 doi:10.1111/j.1744-7909.2008.00754.x.
- 6 Taylor, M., 2018: Climate-smart agriculture: what is it good for? *J. Peasant Stud.*, **45**, 89–107,
7 doi:10.1080/03066150.2017.1312355.
8 <https://www.tandfonline.com/doi/full/10.1080/03066150.2017.1312355> (Accessed October 31,
9 2018).
- 10 Tebaldi, C., and D. Lobell, 2018: Estimated impacts of emission reductions on wheat and maize
11 crops. *Clim. Change*, **146**, 533–545, doi:10.1007/s10584-015-1537-5.
- 12 Teixeira, E. I., J. De Ruiter, A. Ausseil, A. Daigneault, P. Johnstone, A. Holmes, A. Tait, and F.
13 Ewert, 2017: Adapting crop rotations to climate change in regional impact modelling
14 assessments. *Sci. Total Environ.*, **In press**, doi:10.1016/j.scitotenv.2017.10.247.
- 15 Terblanche, J. S., S. Clusella-Trullas, J. A. Deere, and S. L. Chown, 2008: Thermal tolerance in a
16 south-east African population of the tsetse fly *Glossina pallidipes* (Diptera, Glossinidae):
17 Implications for forecasting climate change impacts. *J. Insect Physiol.*, **54**, 114–127,
18 doi:10.1016/J.JINSPHYS.2007.08.007.
19 <https://www.sciencedirect.com/science/article/pii/S002219100700193X> (Accessed March 4,
20 2019).
- 21 Termeer, C. J. A. M., A. Dewulf, G. R. Biesbroek, C. J. A. M. Termeer, A. Dewulf, and G. R.
22 Biesbroek, 2016: Transformational change: governance interventions for climate change
23 adaptation from a continuous change perspective. *J. Environ. Plan. Manag.*, **568**,
24 doi:10.1080/09640568.2016.1168288.
- 25 Termeer, C. J. A. M., A. Dewulf, and G. R. Biesbroek, 2017: Transformational change: governance
26 interventions for climate change adaptation from a continuous change perspective. *J. Environ.*
27 *Plan. Manag.*, **60**, 558–576, doi:10.1080/09640568.2016.1168288.
- 28 Tesfaye, K., and Coauthors, 2017: Climate change impacts and potential benefits of heat-tolerant
29 maize in South Asia. *Theor. Appl. Climatol.*, **130**, 959–970, doi:10.1007/s00704-016-1931-6.
- 30 The Royal Society, 2018: *Greenhouse gas removal*. 1-136 pp. <https://royalsociety.org/-/media/policy/projects/greenhouse-gas-removal/royal-society-greenhouse-gas-removal-report-2018.pdf>
31 (Accessed April 13, 2019).
- 32
- 33 Theurl, M. C., H. Haberl, K.-H. Erb, and T. Lindenthal, 2014: Contrasted greenhouse gas emissions
34 from local versus long-range tomato production. *Agron. Sustain. Dev.*, **34**, 593–602,
35 doi:10.1007/s13593-013-0171-8. <http://link.springer.com/10.1007/s13593-013-0171-8>
36 (Accessed March 17, 2019).
- 37 Thiel, B., and Coauthors, 2015: *Climate Change Adaptation and On-Farm Drainage Management in*
38 *Delta, British Columbia: Current Knowledge and Practices. Final Report for the Delta*
39 *Farmers' Institute, Drainage and Sub-irrigation in Delta, British Columbia Project*. 75 pp.
40 [https://www.bcagclimateaction.ca/wp/wp-content/media/DL09-Delta-Drainage-Sub-irrigation-](https://www.bcagclimateaction.ca/wp/wp-content/media/DL09-Delta-Drainage-Sub-irrigation-full.pdf)
41 [full.pdf](https://www.bcagclimateaction.ca/wp/wp-content/media/DL09-Delta-Drainage-Sub-irrigation-full.pdf).
- 42 Thomas, T. N., M. Leander-Griffith, V. Harp, and J. P. Cioffi, 2015: *National Preparedness Month —*
43 *September 2015 Influences of Preparedness Knowledge and Beliefs on Household Disaster*
44 *Preparedness*. 965-971 pp. <https://www.cdc.gov/mmwr/pdf/wk/mm6435.pdf>.
- 45 Thompson, B., M. J. Cohen, and J. Meerman, 2012: *The Impact of Climate Change and Bioenergy on*
46 *Nutrition*. Springer Netherlands, Dordrecht, 103 pp.
47 http://www.springerlink.com/index/10.1007/978-94-007-0110-6_1 (Accessed December 8,
48 2017).

- 1 Thompson, M. S., 2018: Critical perspectives on gender, food and political economy. *Handbook of*
2 *the political economy of gender*, Edward Elgar Publishing, 470–485.
- 3 Thomson, L. J., S. Macfadyen, and A. A. Hoffmann, 2010: Predicting the effects of climate change on
4 natural enemies of agricultural pests. *Biol. Control*, **52**, 296–306,
5 doi:10.1016/j.biocontrol.2009.01.022.
- 6 Thornton, P. K., J. van de Steeg, A. Notenbaert, and M. Herrero, 2009: The impacts of climate change
7 on livestock and livestock systems in developing countries: A review of what we know and what
8 we need to know. *Agric. Syst.*, **101**, 113–127, doi:10.1016/J.AGSY.2009.05.002.
9 <https://www.sciencedirect.com/science/article/pii/S0308521X09000584> (Accessed March 4,
10 2019).
- 11 Thornton, P. K., P. J. Ericksen, M. Herrero, and A. J. Challinor, 2014: Climate variability and
12 vulnerability to climate change: a review. *Glob. Chang. Biol.*, **20**, 3313–3328,
13 doi:10.1111/gcb.12581. <http://doi.wiley.com/10.1111/gcb.12581> (Accessed March 16, 2019).
- 14 Thu Thuy, P., B. M. Campbell, and S. Garnett, 2009: Lessons for Pro-Poor Payments for
15 Environmental Services: An Analysis of Projects in Vietnam. *Asia Pacific J. Public Adm.*, **31**,
16 117–133, doi:10.1080/23276665.2009.10779359.
17 <http://www.tandfonline.com/doi/abs/10.1080/23276665.2009.10779359> (Accessed April 17,
18 2019).
- 19 Tigchelaar, M., D. S. Battisti, R. L. Naylor, and D. K. Ray, 2018: Future warming increases
20 probability of globally synchronized maize production shocks. *Proc. Natl. Acad. Sci.*, **115**,
21 6644–6649, doi:10.1073/pnas.1718031115.
22 <http://www.pnas.org/lookup/doi/10.1073/pnas.1718031115>.
- 23 Tilman, D., and M. Clark, 2014: Global diets link environmental sustainability and human health.
24 *Nature*, **515**, 518–522, doi:10.1038/nature13959.
25 <http://www.nature.com/doi/10.1038/nature13959>.
- 26 —, C. Balzer, J. Hill, and B. L. Befort, 2011: Global food demand and the sustainable
27 intensification of agriculture. *Proc. Natl. Acad. Sci. U. S. A.*, **108**, 20260–20264,
28 doi:10.1073/pnas.1116437108.
- 29 Tirado, M. C., R. Clarke, L. A. Jaykus, A. McQuatters-Gollop, and J. M. Frank, 2010: Climate change
30 and food safety: A review. *Food Res. Int.*, **43**, 1745–1765,
31 doi:10.1016/J.FOODRES.2010.07.003.
32 <https://www.sciencedirect.com/science/article/pii/S0963996910002231> (Accessed April 23,
33 2018).
- 34 Tirado, M. C., and Coauthors, 2013: Climate Change and Nutrition: Creating a Climate for Nutrition
35 Security. *Food Nutr. Bull.*, **34**, 533–547, doi:10.1177/156482651303400415.
36 <http://journals.sagepub.com/doi/10.1177/156482651303400415> (Accessed April 7, 2019).
- 37 Tiwari, P. C., and B. Joshi, 2012: Natural and socio-economic factors affecting food security in the
38 Himalayas. *Food Secur.*, **4**, 195–207.
- 39 Tobarra, M. A., L. A. López, M.-A. Cadarso, N. Gomez, and I. Cazarro, 2018: Is seasonal
40 households' consumption good for the nexus carbon/water footprint? The Spanish fruits and
41 vegetables case. *Environ. Sci. Technol.*, doi:10.1021/acs.est.8b00221.
42 <https://doi.org/10.1021/acs.est.8b00221>.
- 43 Toensmeier, E., 2016: *The carbon farming solution: a global toolkit of perennial crops and*
44 *regenerative agriculture practices for climate change mitigation and food security*. 480 pp.
45 <https://www.chelseagreen.com/product/the-carbon-farming-solution/> (Accessed March 9, 2019).
- 46 Tollefson, J., 2018: Deforestation ticks up in Brazil's savannah: The Cerrado is the most threatened
47 biome in Brazil, environmentalists proclaim. *Nature*, [https://www.nature.com/articles/d41586-](https://www.nature.com/articles/d41586-018-05695-9)
48 [018-05695-9](https://www.nature.com/articles/d41586-018-05695-9).

- 1 Tontisirin, K., G. Nantel, and L. Bhattacharjee, 2002: Food-based strategies to meet the challenges of
2 micronutrient malnutrition in the developing world. *Proc. Nutr. Soc.*, **61**, 243–250,
3 doi:10.1079/PNS2002155. <http://www.ncbi.nlm.nih.gov/pubmed/12133206> (Accessed October
4 22, 2018).
- 5 Tripathi, A., and A. K. Mishra, 2017: Knowledge and passive adaptation to climate change: An
6 example from Indian farmers. *Clim. Risk Manag.*, **16**, 195–207,
7 doi:10.1016/J.CRM.2016.11.002.
8 <https://www.sciencedirect.com/science/article/pii/S2212096316300250> (Accessed April 8,
9 2019).
- 10 Tripathi, A., D. K. Tripathi, D. K. Chauhan, N. Kumar, and G. S. Singh, 2016: Paradigms of climate
11 change impacts on some major food sources of the world: A review on current knowledge and
12 future prospects. *Agric. Ecosyst. Environ.*, **216**, 356–373, doi:10.1016/j.agee.2015.09.034.
13 <http://dx.doi.org/10.1016/j.agee.2015.09.034>.
- 14 Tripoli, M., and J. Schmidhuber, 2018: *Emerging Opportunities for the Application of Blockchain in*
15 *the Agri-food Industry*. Rome and Geneva, 1-40 pp.
16 <http://www.fao.org/3/ca1335en/CA1335EN.pdf> (Accessed April 17, 2019).
- 17 Troell, M., and Coauthors, 2014: Does aquaculture add resilience to the global food system? *Proc.*
18 *Natl. Acad. Sci.*, **111**, 13257–13263, doi:10.1073/pnas.1404067111.
19 <http://www.pnas.org/lookup/doi/10.1073/pnas.1404067111>.
- 20 Tsilini, V., S. Papantoniou, D.-D. Kolokotsa, and E.-A. Maria, 2015: Urban gardens as a solution to
21 energy poverty and urban heat island. *Sustain. Cities Soc.*, **14**, 323–333,
22 doi:10.1016/J.SCS.2014.08.006.
23 <https://www.sciencedirect.com/science/article/pii/S2210670714000924> (Accessed November 5,
24 2018).
- 25 Tubiello, F. N., 2019: Greenhouse Gas Emissions Due to Agriculture. Elsevier, Inc, p. 10.
- 26 Tubiello, F. N., and Coauthors, 2014: Agriculture, forestry and other land use emissions by sources
27 and removals by sinks. *Rome, Italy*.
- 28 Tubiello, F. N., and Coauthors, 2015: The Contribution of Agriculture, Forestry and other Land Use
29 activities to Global Warming, 1990-2012. *Glob. Chang. Biol.*, **21**, 2655–2660,
30 doi:10.1111/gcb.12865. <http://doi.wiley.com/10.1111/gcb.12865> (Accessed January 9, 2018).
- 31 Tucker, C., 2018: Using environmental imperatives to reduce meat consumption: perspectives from
32 New Zealand. *Kōtuitui New Zeal. J. Soc. Sci. Online*, **13**, 99–110,
33 doi:10.1080/1177083X.2018.1452763.
34 <https://www.tandfonline.com/doi/full/10.1080/1177083X.2018.1452763> (Accessed April 5,
35 2019).
- 36 Tukker, A., R. A. Goldbohm, A. de Koning, M. Verheijden, R. Kleijn, O. Wolf, I. Pérez-Domínguez,
37 and J. M. Rueda-Cantuche, 2011: Environmental impacts of changes to healthier diets in
38 Europe. *Ecol. Econ.*, **70**, 1776–1788, doi:10.1016/J.ECOLECON.2011.05.001.
- 39 Tulchinsky, T. H., 2010: Micronutrient Deficiency Conditions: Global Health Issues. *Public Health*
40 *Rev.*, **32**, 243–255, doi:10.1007/BF03391600. <https://doi.org/10.1007/BF03391600>.
- 41 Tuomisto, H. L., and M. J. de Mattos, 2011: Environmental Impacts of Cultured Meat Production.
42 *Environ. Sci. Technol.*, **45**, 6117–6123, doi:10.1021/es200130u.
43 <http://pubs.acs.org/doi/abs/10.1021/es200130u> (Accessed November 4, 2018).
- 44 Tyler, N. J. C., and Coauthors, 2007: Saami reindeer pastoralism under climate change: Applying a
45 generalized framework for vulnerability studies to a sub-arctic social–ecological system. *Glob.*
46 *Environ. Chang.*, **17**, 191–206, doi:10.1016/J.GLOENVCHA.2006.06.001.
47 <https://www.sciencedirect.com/science/article/abs/pii/S0959378006000550> (Accessed March
48 15, 2019).

- 1 Tylianakis, J. M., R. K. Didham, J. Bascompte, and D. A. Wardle, 2008: Global change and species
2 interactions in terrestrial ecosystems. *Ecol. Lett.*, **11**, 1351–1363, doi:10.1111/j.1461-
3 0248.2008.01250.x.
- 4 U.S. Department of State, 2016: Briefing on announcement of new measures to address the drought in
5 Ethiopia. <https://2009-2017.state.gov/r/pa/prs/ps/2016/03/253954.htm> (Accessed March 20,
6 2019).
- 7 UNCCD, 2017: *Global Land Outlook*. Bonn, Germany,.
- 8 UNDP, 2013: Overview of linkages between gender and climate change. 6.
9 [https://www.undp.org/content/dam/undp/library/gender/Gender and Environment/PB1-AP-
10 Overview-Gender-and-climate-change.pdf](https://www.undp.org/content/dam/undp/library/gender/Gender%20and%20Environment/PB1-AP-Overview-Gender-and-climate-change.pdf) (Accessed May 9, 2018).
- 11 —, 2016: *Climate Information & Early Warning Systems Communications Toolkit UNDP
12 Programme on Climate Information for Resilient Development in Africa*.
- 13 UNFCCC, 2015: The Paris Agreement. http://unfccc.int/paris_agreement/items/9485.php (Accessed
14 December 20, 2017).
- 15 United Nations, 2015: *Transforming our world: the 2030 Agenda for Sustainable Development*. 41
16 pp. [https://sustainabledevelopment.un.org/content/documents/21252030 Agenda for Sustainable
17 Development web.pdf](https://sustainabledevelopment.un.org/content/documents/21252030%20Agenda%20for%20Sustainable%20Development%20web.pdf).
- 18 —, 2017: *Progress towards the Sustainable Development Goals*. 27 pp.
19 [http://srsg.violenceagainstchildren.org/sites/default/files/documents/docs/report_secretary_gener
20 al_progress_agenda2030_2017.pdf](http://srsg.violenceagainstchildren.org/sites/default/files/documents/docs/report_secretary_general_progress_agenda2030_2017.pdf).
- 21 United Nations Statistics Division, 2016: SDG indicators global database.
22 <https://unstats.un.org/sdgs/indicators/database/>.
- 23 Uraguchi, Z. B., 2010: Food price hikes, food security, and gender equality: assessing the roles and
24 vulnerability of women in households of Bangladesh and Ethiopia. *Gen. Dev.*, **18**, 491–501,
25 doi:10.1080/13552074.2010.521992.
26 <https://www.tandfonline.com/doi/full/10.1080/13552074.2010.521992> (Accessed May 3, 2018).
- 27 USAID, 2017: *Ecosystem-based Adaptation and Food Security*. Washington, DC, 8 pp.
28 [https://www.climatelinks.org/sites/default/files/asset/document/2017_USAID_EbA and Food
29 Security.pdf](https://www.climatelinks.org/sites/default/files/asset/document/2017_USAID_EbA%20and%20Food%20Security.pdf) (Accessed March 9, 2019).
- 30 USDA, 2016: A Guide to Smart Snacks in School. 16.
31 <https://healthymeals.fns.usda.gov/sites/default/files/uploads/USDASmartSnacks.pdf>.
- 32 Valdivia, R. O., and Coauthors, 2015: Representative agricultural pathways and scenarios for regional
33 integrated assessment of climate change impacts, vulnerability, and adaptation. *Handbook of
34 Climate Change and Agroecosystems: The Agricultural Model Intercomparison and
35 Improvement Project Integrated Crop and Economic Assessments, Part 1*, World Scientific,
36 101–145.
- 37 Valin, H., and Coauthors, 2014: The future of food demand: understanding differences in global
38 economic models. *Agric. Econ.*, **45**, 51–67, doi:10.1111/agec.12089.
39 <http://doi.wiley.com/10.1111/agec.12089>.
- 40 Vandermeer, J., 1995: The Ecological Basis of Alternative Agriculture. *Annu. Rev. Ecol. Syst.*, **26**,
41 201–224, doi:10.1146/annurev.es.26.110195.001221.
42 <http://www.annualreviews.org/doi/10.1146/annurev.es.26.110195.001221> (Accessed March 4,
43 2019).
- 44 Vandermeer, J., M. van Noordwijk, J. Anderson, C. Ong, and I. Perfecto, 1998: Global change and
45 multi-species agroecosystems: Concepts and issues. *Agric. Ecosyst. Environ.*, **67**, 1–22,
46 doi:10.1016/S0167-8809(97)00150-3.
47 <https://www.sciencedirect.com/science/article/pii/S0167880997001503> (Accessed April 14,
48 2019).

- 1 Vandermeersch, T., R. A. F. Alvarenga, P. Ragaert, and J. Dewulf, 2014: Environmental
2 sustainability assessment of food waste valorization options. *Resour. Conserv. Recycl.*, **87**, 57–
3 64, doi:10.1016/J.RESCONREC.2014.03.008.
4 <https://www.sciencedirect.com/science/article/pii/S0921344914000664?via%3Dihub> (Accessed
5 March 17, 2019).
- 6 Vanham, D., A. Y. Hoekstra, and G. Bidoglio, 2013: Potential water saving through changes in
7 European diets. *Environ. Int.*, **61**, 45–56, doi:10.1016/J.ENVINT.2013.09.011.
8 <https://www.sciencedirect.com/science/article/pii/S0160412013002055> (Accessed March 24,
9 2019).
- 10 Vanlauwe, B., and Coauthors, 2014: Sustainable intensification and the African smallholder farmer.
11 *Curr. Opin. Environ. Sustain.*, **8**, 15–22, doi:10.1016/j.cosust.2014.06.001.
- 12 Vasanth, M., and Coauthors, 2016: Methodological approach for the collection and simultaneous
13 estimation of greenhouse gases emission from aquaculture ponds. *Environ. Monit. Assess.*, **188**,
14 671, doi:10.1007/s10661-016-5646-z. <http://link.springer.com/10.1007/s10661-016-5646-z>
15 (Accessed April 23, 2018).
- 16 Vasconcelos, A. C. F., M. Bonatti, S. L. Schlindwein, L. R. D’Agostini, L. R. Homem, and R. Nelson,
17 2013: Landraces as an adaptation strategy to climate change for smallholders in Santa Catarina,
18 Southern Brazil. *Land use policy*, **34**, 250–254, doi:10.1016/J.LANDUSEPOL.2013.03.017.
- 19 Vaughan, M., D. Backhouse, and E. M. Del Ponte, 2016: Climate change impacts on the ecology of
20 *Fusarium graminearum* species complex and susceptibility of wheat to *Fusarium* head blight: a
21 review. *World Mycotoxin J.*, **9**, 685–700, doi:10.3920/wmj2016.2053.
- 22 Vaughan, N. E., and C. Gough, 2016: Expert assessment concludes negative emissions scenarios may
23 not deliver. *Environ. Res. Lett.*, **11**, 95003, doi:10.1088/1748-9326/11/9/095003.
- 24 —, —, S. Mander, E. W. Littleton, A. Welfle, D. E. H. J. Gernaat, and D. P. van Vuuren, 2018:
25 Evaluating the use of biomass energy with carbon capture and storage in low emission scenarios.
26 *Environ. Res. Lett.*, **13**, 44014.
- 27 Vellakkal, S., J. Fledderjohann, S. Basu, S. Agrawal, S. Ebrahim, O. Campbell, P. Doyle, and D.
28 Stuckler, 2015: Food Price Spikes Are Associated with Increased Malnutrition among Children
29 in Andhra Pradesh, India. *J. Nutr.*, **145**, 1942–1949, doi:10.3945/jn.115.211250.
30 <http://www.ncbi.nlm.nih.gov/pubmed/26136589> (Accessed November 3, 2018).
- 31 Vermeulen, S. J., and Coauthors, 2012: Options for support to agriculture and food security under
32 climate change. *Environ. Sci. Policy*, **15**, 136–144, doi:10.1016/j.envsci.2011.09.003.
33 <http://dx.doi.org/10.1016/j.envsci.2011.09.003>.
- 34 Vermeulen, S. J., and Coauthors, 2013: Addressing uncertainty in adaptation planning for agriculture.
35 *Proc. Natl. Acad. Sci.*, **110**, 8357 LP-8362, doi:10.1073/pnas.1219441110.
36 <http://www.pnas.org/content/110/21/8357.abstract>.
- 37 —, D. Dinesh, S. M. Howden, L. Cramer, and P. K. Thornton, 2018: Transformation in Practice: A
38 Review of Empirical Cases of Transformational Adaptation in Agriculture Under Climate
39 Change. *Front. Sustain. Food Syst.*, **2**, 65.
40 <https://www.frontiersin.org/article/10.3389/fsufs.2018.00065>.
- 41 Vetter, S. H., and Coauthors, 2017: Greenhouse gas emissions from agricultural food production to
42 supply Indian diets: Implications for climate change mitigation. *Agric. Ecosyst. Environ.*, **237**,
43 234–241, doi:10.1016/j.agee.2016.12.024.
- 44 Vieux, F., L.-G. Soler, D. Touazi, and N. Darmon, 2013: High nutritional quality is not associated
45 with low greenhouse gas emissions in self-selected diets of French adults. *Am. J. Clin. Nutr.*, **97**,
46 569–583, doi:10.3945/ajcn.112.035105.
- 47 —, M. Perignon, R. Gazan, and N. Darmon, 2018: Dietary changes needed to improve diet
48 sustainability: are they similar across Europe? *Eur. J. Clin. Nutr.*, **72**, 951–960,

- 1 doi:10.1038/s41430-017-0080-z.
- 2 Vignola, R., C. A. Harvey, P. Bautista-Solis, J. Avelino, B. Rapidel, C. Donatti, and R. Martinez,
3 2015: Ecosystem-based adaptation for smallholder farmers: Definitions, opportunities and
4 constraints. *Agric. Ecosyst. Environ.*, **211**, 126–132, doi:10.1016/J.AGEE.2015.05.013.
5 <https://www.sciencedirect.com/science/article/pii/S0167880915002157> (Accessed March 21,
6 2019).
- 7 Vincent, K., and T. Cull, 2014: Using Indicators to Assess Climate Change Vulnerabilities: Are There
8 Lessons to Learn for Emerging Loss and Damage Debates? *Geogr. Compass*, **8**, 1–12,
9 doi:10.1111/gec3.12105. <http://doi.wiley.com/10.1111/gec3.12105> (Accessed May 31, 2018).
- 10 Virto, I., P. Barré, A. Burlot, and C. Chenu, 2011: Carbon input differences as the main factor
11 explaining the variability in soil organic C storage in no-tilled compared to inversion tilled
12 agrosystems. *Biogeochemistry*, **108**, 17–26, doi:10.1007/s10533-011-9600-4.
- 13 Vranken, L., T. Avermaete, D. Petalios, and E. Mathijs, 2014: Curbing global meat consumption:
14 Emerging evidence of a second nutrition transition. *Environ. Sci. Policy*, **39**, 95–106,
15 doi:10.1016/J.ENVSCI.2014.02.009.
- 16 van Vuuren, D. P., and Coauthors, 2014: A new scenario framework for Climate Change Research:
17 Scenario matrix architecture. *Clim. Change*, **122**, 373–386, doi:10.1007/s10584-013-0906-1.
- 18 —, and Coauthors, 2017: Energy, land-use and greenhouse gas emissions trajectories under a green
19 growth paradigm. *Glob. Environ. Chang.*, **42**, 237–250, doi:10.1016/j.gloenvcha.2016.05.008.
20 <http://dx.doi.org/10.1016/j.gloenvcha.2016.05.008>.
- 21 —, and Coauthors, 2018: Alternative pathways to the 1.5 °C target reduce the need for negative
22 emission technologies. *Nat. Clim. Chang.*, 1–7, doi:10.1038/s41558-018-0119-8.
- 23 Waha, K., C. Müller, A. Bondeau, J. P. Dietrich, P. Kurukulasuriya, J. Heinke, and H. Lotze-Campen,
24 2013: Adaptation to climate change through the choice of cropping system and sowing date in
25 sub-Saharan Africa. *Glob. Environ. Chang.*, **23**, 130–143, doi:10.1016/j.gloenvcha.2012.11.001.
- 26 Waha, K., and Coauthors, 2017: Climate change impacts in the Middle East and Northern Africa
27 (MENA) region and their implications for vulnerable population groups. *Reg. Environ. Chang.*,
28 **17**, 1623–1638, doi:10.1007/s10113-017-1144-2.
29 <https://link.springer.com/content/pdf/10.1007%2Fs10113-017-1144-2.pdf> (Accessed February
30 27, 2019).
- 31 Wakeland, W., S. Cholette, and K. Venkat, 2012: Green Technologies in Food Production and
32 Processing. 211–236, doi:10.1007/978-1-4614-1587-9.
- 33 Waldner, F., S. Fritz, A. Di Gregorio, and P. Defourny, 2015: Mapping priorities to focus cropland
34 mapping activities: Fitness assessment of existing global, regional and national cropland maps.
35 *Remote Sens.*, **7**, 7959–7986, doi:10.3390/rs70607959.
- 36 Waldron, A., D. Garrity, Y. Malhi, C. Girardin, D. C. Miller, and N. Seddon, 2017: Agroforestry Can
37 Enhance Food Security While Meeting Other Sustainable Development Goals. *Trop. Conserv.
38 Sci.*, **10**, 194008291772066, doi:10.1177/1940082917720667.
- 39 Walther, G.-R., and Coauthors, 2009: Alien species in a warmer world: risks and opportunities.
40 *Trends Ecol. Evol.*, **24**, 686–693, doi:10.1016/j.tree.2009.06.008.
41 <http://www.ncbi.nlm.nih.gov/pubmed/19712994> (Accessed October 29, 2018).
- 42 Wang, C. Y., H. W. Huang, C. P. Hsu, and B. B. Yang, 2016: Recent Advances in Food Processing
43 Using High Hydrostatic Pressure Technology. *Crit. Rev. Food Sci. Nutr.*, **56**, 527–540,
44 doi:10.1080/10408398.2012.745479.
- 45 Wang, P., Z. Zhang, X. Song, Y. Chen, X. Wei, P. Shi, and F. Tao, 2014: Temperature variations and
46 rice yields in China: Historical contributions and future trends. *Clim. Change*, **124**, 777–789,
47 doi:10.1007/s10584-014-1136-x.

- 1 Waongo, M., P. Laux, and H. Kunstmann, 2015: Adaptation to climate change: The impacts of
2 optimized planting dates on attainable maize yields under rainfed conditions in Burkina Faso.
3 *Agric. For. Meteorol.*, **205**, 23–39, doi:10.1016/j.agrformet.2015.02.006.
4 <http://dx.doi.org/10.1016/j.agrformet.2015.02.006>.
- 5 Ward, F. A., S. A. Amer, and F. Ziaee, 2013: Water allocation rules in Afghanistan for improved food
6 security. *Food Secur.*, **5**, 35–53.
- 7 Warner, B. P., C. Kuzdas, M. G. Yglesias, and D. L. Childers, 2015: Limits to adaptation to
8 interacting global change risks among smallholder rice farmers in Northwest Costa Rica. *Glob.*
9 *Environ. Chang.*, doi:10.1016/j.gloenvcha.2014.11.002.
- 10 Warner, K., C. Ehrhart, A. de Sherbinin, S. Adamo, and T. Chai-Onn, 2009: In search of shelter.
11 *Mapp. Eff. Clim. Chang. Hum. Migr. Displac. Bonn CARE*,.
- 12 —, T. Afifi, K. Henry, T. Rawe, C. Smith, and A. De Sherbinin, 2012: Where the rain falls:
13 Climate change, food and livelihood security, and migration. *Glob. Policy Rep. Where Rain*
14 *Falls Proj. Bonn CARE Fr. UNU-EHS*,.
- 15 Warren, R., J. Price, E. Graham, N. Forstenhaeusler, and J. VanDerWal, 2018: The projected effect on
16 insects, vertebrates, and plants of limiting global warming to 1.5°C rather than 2°C. *Science* (80-
17 .), **360**, 791 LP-795. <http://science.sciencemag.org/content/360/6390/791.abstract>.
- 18 Waryszak, P., T. I. Lenz, M. R. Leishman, and P. O. Downey, 2018: Herbicide effectiveness in
19 controlling invasive plants under elevated CO₂: Sufficient evidence to rethink weeds
20 management. *J. Environ. Manage.*, **226**, 400–407, doi:10.1016/J.JENVMAN.2018.08.050.
21 <https://www.sciencedirect.com/science/article/pii/S0301479718309216> (Accessed March 4,
22 2019).
- 23 Watanabe, M., H. Shiogama, Y. Imada, M. Mori, M. Ishii, and M. Kimoto, 2013: Event Attribution of
24 the August 2010 Russian Heat Wave. *SOLA*, **9**, 65–68.
- 25 Wattnem, T., 2016: Seed laws, certification and standardization: outlawing informal seed systems in
26 the Global South. *J. Peasant Stud.*, **43**, 850–867, doi:10.1080/03066150.2015.1130702.
- 27 Watts, N., and Coauthors, 2018: The Lancet Countdown on health and climate change: from 25 years
28 of inaction to a global transformation for public health. *Lancet*, **391**, 581–630,
29 doi:[https://doi.org/10.1016/S0140-6736\(17\)32464-9](https://doi.org/10.1016/S0140-6736(17)32464-9).
30 <http://www.sciencedirect.com/science/article/pii/S0140673617324649>.
- 31 Weber, C. L., and H. S. Matthews, 2008: Food-Miles and the Relative Climate Impacts of Food
32 Choices in the United States. *Environ. Sci. Technol.*, **42**, 3508–3513, doi:10.1021/es702969f.
33 <http://pubs.acs.org/doi/abs/10.1021/es702969f> (Accessed May 31, 2018).
- 34 WEF/McKinsey & Company, 2018: *Innovation with a Purpose: The role of technology innovation in*
35 *accelerating food systems transformation*. Geneva, Switzerland, 42 pp.
36 http://www3.weforum.org/docs/WEF_Innovation_with_a_Purpose_VF-reduced.pdf (Accessed
37 May 18, 2018).
- 38 Wegren, S. K., 2011: Food Security and Russia's 2010 Drought. *Eurasian Geogr. Econ.*, **52**, 140–
39 156, doi:10.2747/1539-7216.52.1.140. [http://www.tandfonline.com/doi/abs/10.2747/1539-
40 7216.52.1.140](http://www.tandfonline.com/doi/abs/10.2747/1539-7216.52.1.140).
- 41 Weiss, F., and A. Leip, 2012: Greenhouse gas emissions from the EU livestock sector: A life cycle
42 assessment carried out with the CAPRI model. *Agric. Ecosyst. Environ.*, **149**, 124–134,
43 doi:10.1016/J.AGEE.2011.12.015.
44 <https://www.sciencedirect.com/science/article/pii/S0167880911004415> (Accessed May 25,
45 2018).
- 46 Wellesley, L., C. Happer, and A. Froggatt, 2015: *Chatham House Report Changing Climate,*
47 *Changing Diets Pathways to Lower Meat Consumption*. www.chathamhouse.org (Accessed
48 April 24, 2019).

- 1 —, F. Preston, J. Lehne, and R. Bailey, 2017: Chokepoints in global food trade: Assessing the risk.
2 *Res. Transp. Bus. Manag.*, **25**, 15–28, doi:10.1016/J.RTBM.2017.07.007.
3 <https://www.sciencedirect.com/science/article/pii/S2210539517300172> (Accessed January 8,
4 2018).
- 5 Welton, G., 2011: The impact of Russia’s 2010 grain export ban. *Oxfam Policy Pract. Agric. Food L.*,
6 **11**, 76–107.
- 7 Werrell, C. E., F. Femia, and T. Sternberg, 2015: Did We See It Coming?: State Fragility, Climate
8 Vulnerability, and the Uprisings in Syria and Egypt. *SAIS Rev. Int. Aff.*, **35**, 29–46,
9 doi:10.1353/sais.2015.0002.
10 https://muse.jhu.edu/content/crossref/journals/sais_review/v035/35.1.werrell.html (Accessed
11 May 22, 2018).
- 12 Westendorf, M. L., 2000: Food Waste as Animal Feed: an Introduction. *Food Waste to Animal Feed*,
13 Iowa State University Press, Ames, Iowa, USA, 3–16
14 <http://doi.wiley.com/10.1002/9780470290217.ch1> (Accessed April 17, 2019).
- 15 Westhoek, H., and Coauthors, 2014: Food choices, health and environment: Effects of cutting
16 Europe’s meat and dairy intake. *Glob. Environ. Chang.*, **26**, 196–205,
17 doi:10.1016/j.gloenvcha.2014.02.004.
18 <http://www.sciencedirect.com/science/article/pii/S0959378014000338> (Accessed January 11,
19 2018).
- 20 Wezel, A., G. Soboksa, S. McClelland, F. Delespesse, and A. Boissau, 2015: The blurred boundaries
21 of ecological, sustainable, and agroecological intensification: a review. *Agron. Sustain. Dev.*, **35**,
22 1283–1295, doi:10.1007/s13593-015-0333-y. [http://link.springer.com/10.1007/s13593-015-](http://link.springer.com/10.1007/s13593-015-0333-y)
23 [0333-y](http://link.springer.com/10.1007/s13593-015-0333-y) (Accessed March 3, 2019).
- 24 WFP, 2017: *At the root of Exodus: Food security, conflict and international migration*. 1-74 pp.
25 [http://documents.wfp.org/stellent/groups/public/documents/communications/wfp291884.pdf?_g](http://documents.wfp.org/stellent/groups/public/documents/communications/wfp291884.pdf?_ga=2.235970817.1207453463.1495448038-536181563.1495448038%0Awww.wfp.org/at-the-root-of-exodus)
26 [a=2.235970817.1207453463.1495448038-536181563.1495448038%0Awww.wfp.org/at-the-](http://documents.wfp.org/stellent/groups/public/documents/communications/wfp291884.pdf?_ga=2.235970817.1207453463.1495448038-536181563.1495448038%0Awww.wfp.org/at-the-root-of-exodus)
27 [root-of-exodus](http://documents.wfp.org/stellent/groups/public/documents/communications/wfp291884.pdf?_ga=2.235970817.1207453463.1495448038-536181563.1495448038%0Awww.wfp.org/at-the-root-of-exodus).
- 28 —, 2018: Climate action | World Food Programme. <http://www1.wfp.org/climate-action> (Accessed
29 November 4, 2018).
- 30 Whitfield, A., S. Bergmann, and J. Lazarchick, 2015: *Iron Deficiency Anemia Diagnosed in Female*
31 *Teenagers*. <https://www.jscimedcentral.com/FamilyMedicine/familymedicine-2-1058.pdf>
32 (Accessed October 8, 2018).
- 33 WHO, 2015: Obesity and overweight. *World Heal. Organ.*, 1–5.
34 <http://www.who.int/mediacentre/factsheets/fs311/en/>.
- 35 Wiebe, K., and Coauthors, 2015: Climate change impacts on agriculture in 2050 under a range of
36 plausible socioeconomic and emissions scenarios. *Environ. Res. Lett.*, **10**, 85010,
37 doi:10.1088/1748-9326/10/8/085010. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b)
38 [9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b](http://stacks.iop.org/1748-9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b).
- 39 Wilhelm, J., and R. Smith, 2018: Ecosystem services and land sparing potential of urban and peri-
40 urban agriculture: A review. *Renew. Agric. Food Syst.*, **33**, 481–494.
- 41 Willett, W., and Coauthors, 2019: Food in the Anthropocene: the EAT-Lancet Commission on healthy
42 diets from sustainable food systems. *Lancet (London, England)*, **393**, 447–492,
43 doi:10.1016/S0140-6736(18)31788-4. <http://www.ncbi.nlm.nih.gov/pubmed/30660336>
44 (Accessed March 4, 2019).
- 45 Williams, J., and P. J. Crutzen, 2010: Nitrous oxide from aquaculture. *Nat. Geosci.*, **3**, 143–143,
46 doi:10.1038/ngeo804. <http://www.nature.com/articles/ngeo804> (Accessed April 23, 2018).
- 47 Wirsenius, S., F. Hedenus, and K. Mohlin, 2011: Greenhouse gas taxes on animal food products:
48 rationale, tax scheme and climate mitigation effects. *Clim. Change*, **108**, 159–184,

- 1 doi:10.1007/s10584-010-9971-x.
- 2 Wittman, H., 2011: Food sovereignty: a new rights framework for food and nature? *Environ. Soc.*, **2**,
3 87–105.
- 4 Wolff, S., E. A. Schrammeijer, C. J. E. Schulp, and P. H. Verburg, 2018: Meeting global land
5 restoration and protection targets: What would the world look like in 2050? *Glob. Environ.*
6 *Chang.*, **52**, 259–272, doi:10.1016/J.GLOENVCHA.2018.08.002.
7 <https://www.sciencedirect.com/science/article/abs/pii/S0959378017311718> (Accessed March
8 11, 2019).
- 9 Wollenberg, E., and Coauthors, 2016: Reducing emissions from agriculture to meet the 2 °C target.
10 *Glob. Chang. Biol.*, **22**, 1–6, doi:10.1111/gcb.13340.
- 11 Wood, S. A., D. Tirfessa, and F. Baudron, 2018: Soil organic matter underlies crop nutritional quality
12 and productivity in smallholder agriculture. *Agric. Ecosyst. Environ.*, **266**, 100–108,
13 doi:10.1016/j.agee.2018.07.025.
14 <https://linkinghub.elsevier.com/retrieve/pii/S0167880918303025>.
- 15 Woolf, D., J. E. Amonette, F. A. Street-Perrott, J. Lehmann, and S. Joseph, 2010: Sustainable biochar
16 to mitigate global climate change. *Nat. Commun.*, **1**, 1–9, doi:10.1038/ncomms1053.
17 <http://www.nature.com/doi/10.1038/ncomms1053> (Accessed April 25, 2019).
- 18 World Bank, 2015: The cost of the gender gap in agricultural productivity in Malawi, Tanzania, and
19 Uganda. 1–42. [http://documents.worldbank.org/curated/en/847131467987832287/The-cost-of-](http://documents.worldbank.org/curated/en/847131467987832287/The-cost-of-the-gender-gap-in-agricultural-productivity-in-Malawi-Tanzania-and-Uganda)
20 the-gender-gap-in-agricultural-productivity-in-Malawi-Tanzania-and-Uganda (Accessed
21 November 4, 2018).
- 22 —, 2019: Agriculture, forestry, and fishing, value added (% of GDP) | Data.
23 https://data.worldbank.org/indicator/nv.agr.totl.zs?year_high_desc=true (Accessed April 15,
24 2019).
- 25 Yan, Y. W., Y. C. Wang, C. C. Feng, P. H. M. Wan, and K. T. T. Chang, 2017: Potential
26 distributional changes of invasive crop pest species associated with global climate change. *Appl.*
27 *Geogr.*, **82**, 83–92, doi:10.1016/j.apgeog.2017.03.011.
- 28 Yang, P., Q. He, J. Huang, and C. Tong, 2015: Fluxes of greenhouse gases at two different
29 aquaculture ponds in the coastal zone of southeastern China. *Atmos. Environ.*, **115**, 269–277,
30 doi:10.1016/J.ATMOSENV.2015.05.067.
31 <https://www.sciencedirect.com/science/article/pii/S1352231015301382> (Accessed April 23,
32 2018).
- 33 Yao, Z., L. Zhang, S. Tang, X. Li, and T. Hao, 2017: The basic characteristics and spatial patterns of
34 global cultivated land change since the 1980s. *J. Geogr. Sci.*, **27**, 771–785, doi:10.1007/s11442-
35 017-1405-5. <http://link.springer.com/10.1007/s11442-017-1405-5> (Accessed September 18,
36 2018).
- 37 Yeh, E. T., Y. Nyima, K. A. Hopping, and J. A. Klein, 2014: Tibetan Pastoralists’ Vulnerability to
38 Climate Change: A Political Ecology Analysis of Snowstorm Coping Capacity. *Hum. Ecol.*, **42**,
39 61–74, doi:10.1007/s10745-013-9625-5. <http://link.springer.com/10.1007/s10745-013-9625-5>
40 (Accessed March 15, 2019).
- 41 Yu, Y. Q., W. Zhang, and Y. Huang, 2014: Impact assessment of climate change, carbon dioxide
42 fertilization and constant growing season on rice yields in China. *Clim. Change*, **124**, 763–775,
43 doi:10.1007/s10584-014-1129-9.
- 44 van Zanten, H. H. E., and Coauthors, 2018: Defining the planets land boundary for sustainable
45 livestock consumption. *Glob. Chang. Biol. Press.*,
- 46 Zech, K. M., and U. A. Schneider, 2019: Carbon leakage and limited efficiency of greenhouse gas
47 taxes on food products. *J. Clean. Prod.*, **213**, 99–103, doi:10.1016/J.JCLEPRO.2018.12.139.
- 48 Zhang, G., and Coauthors, 2017: Spatiotemporal patterns of paddy rice croplands in China and India

- 1 from 2000 to 2015. *Sci. Total Environ.*, **579**, 82–92, doi:10.1016/j.scitotenv.2016.10.223.
- 2 Zhao, C., S. Piao, Y. Huang, X. Wang, P. Ciais, M. Huang, Z. Zeng, and S. Peng, 2016: Field
3 warming experiments shed light on the wheat yield response to temperature in China. *Nat.*
4 *Commun.*, **7**, 13530, doi:10.1038/ncomms13530. <http://www.nature.com/articles/ncomms13530>
5 (Accessed March 7, 2019).
- 6 —, and Coauthors, 2017a: Temperature increase reduces global yields of major crops in four
7 independent estimates. *Proc. Natl. Acad. Sci. U. S. A.*, **114**, 9326–9331,
8 doi:10.1073/pnas.1701762114. <http://www.ncbi.nlm.nih.gov/pubmed/28811375> (Accessed
9 October 1, 2018).
- 10 Zhao, Z. Q., Z. H. Bai, W. Winiwarter, G. Kieseewetter, C. Heyes, and L. Ma, 2017b: Mitigating
11 ammonia emission from agriculture reduces PM2.5 pollution in the Hai River Basin in China.
12 *Sci. Total Environ.*, **609**, 1152–1160, doi:10.1016/j.scitotenv.2017.07.240.
- 13 Zhu, C., and Coauthors, 2018: Carbon dioxide (CO₂) levels this century will alter the protein,
14 micronutrients, and vitamin content of rice grains with potential health consequences for the
15 poorest rice-dependent countries. *Sci. Adv.*, **4**.
16 <http://advances.sciencemag.org/content/4/5/eaq1012.abstract>.
- 17 Zhu, X., R. Clements, J. Hagggar, A. Quezada, and J. Torres, 2011: *Technologies for Climate Change*
18 *Adaptation - Agriculture Sector*. 1-219 pp.
- 19 Zia, A., and C. H. Wagner, 2015: Mainstreaming Early Warning Systems in Development and
20 Planning Processes: Multilevel Implementation of Sendai Framework in Indus and Sahel. *Int. J.*
21 *Disaster Risk Sci.*, **6**, 189–199, doi:10.1007/s13753-015-0048-3.
- 22 Zilberman, D., L. Lipper, N. McCarthy, and B. Gordon, 2018: Innovation in Response to Climate
23 Change. *Climate Smart Agriculture, Management and Policy*, FAO, 49–74
24 http://link.springer.com/10.1007/978-3-319-61194-5_4.
- 25 Zimmermann, A., H. Webber, G. Zhao, F. Ewert, J. Kros, J. Wolf, W. Britz, and W. De Vries, 2017:
26 Climate change impacts on crop yields, land use and environment in response to crop sowing
27 dates and thermal time requirements. *Agric. Syst.*, **157**, 81–92, doi:10.1016/j.agsy.2017.07.007.
- 28 Ziska, L., and Coauthors, 2018: Climate Change, Carbon Dioxide, and Pest Biology, Managing the
29 Future: Coffee as a Case Study. *Agronomy*, **8**, 152, doi:10.3390/agronomy8080152.
30 <http://www.mdpi.com/2073-4395/8/8/152> (Accessed March 4, 2019).
- 31 Ziska, L. H., and Coauthors, 2016: Rising atmospheric CO₂ is reducing the protein concentration of a
32 floral pollen source essential for North American bees. *Proc. R. Soc. B Biol. Sci.*, **283**,
33 20160414, doi:10.1098/rspb.2016.0414.
34 <http://rspb.royalsocietypublishing.org/lookup/doi/10.1098/rspb.2016.0414> (Accessed March 4,
35 2019).
- 36 Zomer, R. J., H. Neufeldt, J. Xu, A. Ahrends, D. Bossio, A. Trabucco, M. Van Noordwijk, and M.
37 Wang, 2016: Global Tree Cover and Biomass Carbon on Agricultural Land: The contribution of
38 agroforestry to global and national carbon budgets. *Sci. Rep.*, **6**, 29987.
- 39 —, D. A. Bossio, R. Sommer, and L. V Verchot, 2017: Global sequestration potential of increased
40 organic carbon in cropland soils. *Sci. Rep.*, **7**, 15554.
- 41 zu Ermgassen, E. K. H. J., B. Phalan, R. E. Green, and A. Balmford, 2016: Reducing the land use of
42 EU pork production: where there's swill, there's a way. *Food Policy*, **58**, 35–48,
43 doi:10.1016/J.FOODPOL.2015.11.001.
44 <https://www.sciencedirect.com/science/article/pii/S0306919215001256?via%3Dihub> (Accessed
45 April 17, 2019).
- 46 Zubair, L., and Coauthors, 2015: Climate Change Impacts on Rice Farming Systems in Northwestern
47 Sri Lanka. *Handbook of Climate Change and agroecosystems: The Agricultural Model*
48 *Intercomparison and Improvement Project Integrated Crop and Economic Assessments, Part 2*,

- 1 315–352.
- 2
- 3

1 **Supplementary Material**2 **Section SM5.1**

3 **Table SM5.1 A gendered approach to understanding how climate change affects dimensions of food**
 4 **security across pastoral and agro-pastoral livestock-holders (adapted from McKune et al. (2015); Ongoro**
 5 **and Ogara (2012) and Fratkin et al. (2004). ↑ increased, ↓ decreased**

| Group | Livelihoods | Health | Nutrition |
|----------------------|---|---|---|
| <i>Pastoral</i> | <p>↑ time demand on <i>women</i> and <i>girls</i> for water, fuel collection</p> <p>↑ time demand on <i>men</i> to seek out water sources with herd</p> <p>↑ <i>men</i> exposure to attacks from other groups</p> <p>↑ <i>men</i> migration resulting in ↑ <i>women</i> workload</p> <p>↑ productive and reproductive demands on <i>women</i></p> <p>↓ financial autonomy of <i>women</i> due to liquidation of small animal assets</p> <p>↑ women poverty due to livestock losses of men</p> | <p>↑ disease risk due to proximity of <i>women</i>'s work to disease agents</p> <p>↑ <i>children</i> health and growth due to reduced milk consumption</p> <p>↑ <i>women</i> and <i>girls</i> exposure to insecurity and dangers when looking for water</p> <p>↑ <i>women</i> and <i>children</i> vulnerability to water-borne diseases</p> <p>↑ vulnerability to <i>maternal mortality</i> due to ↑ fertility due to sedentarisation</p> <p>↓ mental and emotional health due to increased stress/loss of social support for both <i>men</i> and <i>women</i></p> <p>↑ vulnerability of newly sedentarized households, particularly <i>women</i></p> | <p>↑ undernutrition of <i>men</i> and <i>women</i> due to ↓ availability of plant and animal foods</p> <p>↑ undernutrition of <i>men</i> and <i>women</i> due to separation of from milk-producing animals</p> <p>↑ undernutrition in <i>men</i> and <i>women</i> due to unfavorable trade-offs in diet between animal products and grains</p> <p>↑ risk of food insecurity I <i>men</i> and <i>women</i> due to ↓ production of livestock and ↑ prices</p> |
| <i>Agro-pastoral</i> | <p>↑ time demand on <i>women</i> due to migration of men for herding or wage labor</p> <p>↓ financial autonomy of <i>women</i> due to liquidation of small animal assets</p> <p>↑ constraints on herd management due to shifts in responsibilities</p> <p>↑ susceptibility to market</p> | <p>Earlier weaning, shortened birth intervals, and risk of <i>maternal depletion</i></p> <p>↑ incidence of anemia and stunting in <i>children</i></p> <p>↑ susceptibility to infectious diseases that are sensitive to climate change in both <i>men</i> and <i>women</i></p> <p>↑ <i>child</i> mortality rates</p> | <p>↑ exposure of <i>men</i> and <i>women</i> to foods that have become spoiled</p> <p>Less varied and less nutritious diets for <i>men</i> and <i>women</i></p> <p>↑ malnutrition, including overnutrition, in <i>men</i> and <i>women</i></p> |

| | | | |
|--|--------------|--|--|
| | fluctuations | | |
|--|--------------|--|--|

1

2

1 **Section SM5.2**2 **Table SM5.2 Impacts of selected climate drivers on food security pillars.**

| Food security pillar | Driver of climate change | Process | Impact | Reference |
|-----------------------------|--|--|--|--|
| Availability | Increase in temperature | <ul style="list-style-type: none"> • Increased water demand • Increased heat and drought stress • Shorter growing period • More frequent heat wave • Terminal heat • Reduced grain filling period • Decreased soil fertility • Land degradation • Higher pre-harvest loss due to disease and pest attack • Negative effects on physiological processes | Decreased crop yield and animal performance | Zhao et al. (2017) Asseng et al. (2015) Myers et al. (2017) Ovalle-Rivera et al. (2015) Rosenzweig et al. (2014) Medina et al. (2017) Paterson and Lima (2011) Schlenker and Roberts (2009) |
| | CO ₂ concentration | <ul style="list-style-type: none"> • Increased photosynthesis in C3 crops • Increased water use efficiency | Increased crop yield | Franzaring et al. (2013) Mishra and Agrawal (2014) Myers et al. (2014) Ishigooka et al. (2017) Zhu et al. (2018) Loladze (2014) Yu et al. (2014) |
| | Precipitation (untimely, erratic, decreased) | <ul style="list-style-type: none"> • Drought and heat stress • Crop failure • Land degradation • Reduced soil fertility | Decreased crop yield and pasture stocking rates and animal performance | Leng and Hall (2019) Zscheischler et al. (2018) Meng et al. (2016) Zimmerman et al. (2017) FAO et al. (2018) |

Subject to Copy-editing

| | | | | |
|-------------|---|--|--|---|
| | Extreme events (drought, flood, cyclones etc.) | <ul style="list-style-type: none"> • Decrease in organic matter • Soil erosion • Crop failure • Disruption of distribution and exchange | Decreased crop yield Increased livestock mortality Decreased distribution and exchange | Leng and Hall (2019) Rivera-Ferre (2014) |
| Access | Increase in Temperature | <ul style="list-style-type: none"> • Increase in price • Loss of agricultural income • Disproportionate impact on low-income consumers | Increased food price and reduced purchasing power | Morris et al. (2017) Vermeulen et al. (2012) Abid et al. (2016) Harvey et al. (2014) UNCCD (2017) |
| | Precipitation (untimely, erratic, decreased) | <ul style="list-style-type: none"> • Low yield, price increase • Loss of agricultural income due to reduced yield and productivity • Decrease in barley yield • Inability to invest in adaptation and diversification measures to endure price rises | Increased food price and reduced purchasing power | FAO (2016) Kelley et al. (2015) Morris et al. (2017) Vermeulen et al. (2012) Abid et al. (2016) Harvey et al. (2014) UNCCD (2017) |
| | Extreme Events (drought, flood, cyclones etc.) | <ul style="list-style-type: none"> • Price increase due to low yield or sporadic crop failure • Loss of agricultural income | Increased food price and reduced purchasing power | Valin et al. (2014) Robinson et al. (2014) Nelson et al. (2013) Schmitz et al. (2014) |
| Utilization | Increase in Temperature | <ul style="list-style-type: none"> • Decreased in nutritional content • Increased mycotoxins • Reduced water quantity and | Reduced quality | Tirado and Meerman (2012) Aberman and Tirado (2014) Thompson et al. (2012) |

| | | | | |
|-----------|---|--|---|--|
| | | <p>quality to prepare food</p> <ul style="list-style-type: none"> • Negative impact on food safety • Higher post-harvest loss both in quantity and quality | | |
| | CO ₂ Concentration | <ul style="list-style-type: none"> • Decreased protein content • Less zinc content • Less iron content • Increased biomass but reduced multiple nutrients • Less radiation interception and less biomass production | Reduced quality | <p>Myers et al. (2014)</p> <p>Smith et al. (2017)</p> <p>Myers et al. (2015)</p> <p>Medek et al. (2017)</p> <p>Bahrami et al. (2017)</p> <p>Rosenzweig and Hillel (2015)</p> |
| | Extreme Events (drought, flood, cyclones etc.) | <ul style="list-style-type: none"> • Adverse weather affects food storage and distribution | Reduced quality | <p>Wellesley et al. (2017)</p> <p>Thompson et al. (2012)</p> |
| Stability | Increase in Temperature | <ul style="list-style-type: none"> • Disruption of food supply | Fluctuation in production, supply and price | <p>Allen et al. (2017)</p> <p>Tigchelaar et al. (2018)</p> |
| | Precipitation (untimely, erratic, decreased) | <ul style="list-style-type: none"> • Disruption of food supply • Yield variability • Fluctuation in yield, supply and price • Crop failure due to extreme drought | Fluctuation in production, supply and price | <p>Schmidhuber and Tubiello (2007)</p> <p>Kelley et al. (2015)</p> <p>Selby et al. (2017)</p> <p>Kelley et al. (2017)</p> <p>Medina-Elizalde and Rohling (2012)</p> |

| | | | | |
|--|---|---|---|--|
| | Extreme Events (drought, flood, cyclones etc.) | <ul style="list-style-type: none"> • Impacts on world market export prices that carry through to domestic consumer prices • Widespread crop failure contributing to migration and conflict • Disruption of food supply due to civil disturbance and social tension | Fluctuation in production, supply and price | Kelley et al. (2015) Willenbockel (2012) Hendrix (2018) Selby et al. (2017) Kelley et al. (2017) |
|--|---|---|---|--|

1

2

3 ***Detection and attribution methods***

4 Observed impacts of climate change on food security have been noted as a cause of concern (HLPE
5 2012) and assessed in AR5 (Porter et al. 2014; Cramer et al. 2014) and SR15 (IPCC 2018). Assessing
6 evidence for detection and attribution of observed climate change impacts on the food system remains
7 a challenge because agriculture is a managed system with practices changing over time. Using AR5
8 and SR15 findings that observed climate changes attributable to human influence include rising
9 temperatures, increases in the intensity and frequency of hot days and nights, more areas with
10 increases than decreases in the frequency, intensity, and or amount of heavy precipitation, and drying
11 trends in some regions especially in the Mediterranean region (including southern Europe, northern
12 Africa and the Near East), we assess recent studies of observed climate change impacts on the food
13 system that utilise IPCC attribution methods (Hegerl et al. 2010), as well as others that depend on
14 local knowledge from the developing world.

15 New work has addressed observed climate effects on expanded aspects of the food system, including
16 pastoral systems (Rasul et al. 2019; Abiona et al. 2016), pests, diseases, and pollinators (Bebber et al.
17 2014; Schweiger et al. 2010), and adaptation (Li et al. 2017) (see Section 5.3). Surveys of farmer
18 perceptions of climate changes and their impacts are being increasingly utilised in developing
19 countries for example (Hussain et al. 2016) (Ifeanyi-obi et al. 2016; Onyeneke 2018).

20

21 ***Improvements in projection methods since AR5***

22 Since AR5, methods for assessment of future climate change impacts on food systems have improved
23 in several areas, providing new insights. These methods include greater number of ensembles of
24 multiple climate, crop, and economic models, with improved characterisation of uncertainty (Wiebe et
25 al. 2015); further comparison of results from process-based crop models and statistical models (Zhao
26 et al. 2017); advances in regional integrated assessments (Rosenzweig and Hillel 2015), and new
27 coordinated global and regional studies (Rosenzweig et al. 2017; Ruane et al. 2018). Temperature
28 response functions in crop models have been improved (Wang et al. 2017).

1 Expanded meta-analyses of free-air carbon dioxide experiments (FACE) have examined effects of
2 high CO₂ on crop nutrients not just on yield (Smith and Myers 2018; Zhu et al. 2018) (Section
3 5.2.4.2). Recent reviews have confirmed that higher CO₂ concentrations increase crop growth and
4 yield, especially in crops with C3 photosynthetic pathways, but realisation of these direct CO₂ effects
5 depends on nutrient and water availability (Lombardozi et al. 2018; Toreti et al.; Uddin et al. 2018)
6 (*high confidence*). New work has considered future impacts of farming systems, extreme events, fruits
7 and vegetables, rangelands and livestock, and aquaculture, as well as food safety, pests and diseases,
8 and food quality (Section 5.2).

9 However, several sources of uncertainty exist in projection of climate change crop impacts, partly
10 stemming from differences between the models and methods utilised, sparse observations related to
11 current climate trends, and other agro-ecosystem responses (e.g., to CO₂ effects) (Mistry et al. 2017;
12 Li et al. 2015; Bassu et al. 2014; Asseng et al. 2013). The uncertainty in climate simulations is
13 generally larger than, or sometimes comparable to, the uncertainty in crop simulations using a single
14 model (Iizumi et al. 2011), but is less than crop model uncertainty when multiple crop models are
15 used as in AgMIP (Rosenzweig et al. 2014b) and CO₂ is considered (Hasegawa et al. 2018; Müller et
16 al. 2014; Asseng et al. 2013).

17 Most of the work on projected impacts on climate change impacts on crops continues to focus on the
18 major commodities-wheat, maize, rice, and soybean-while areas still lagging are multi-model
19 ensemble approaches for livestock and fruits and vegetables. While the current reliance on the four
20 major commodities makes assessment of climate change impacts on them important, there is a
21 growing recognition that more than caloric intake is required to achieve food security for all and that
22 assessments need to take into account how climate change will affect the 2 billion malnourished
23 people in the current climate and food system.

24

25

26

27

28

29

30

31

32

33

34

35

36

37

1
2
3**Table SM5.3 Observed climate change impacts on crop production, data sources, and detection and attribution methods**

| Climate observations | Climate data source | Observed impacts | Impact method/source | Time period | Region | Detection & Attribution method | Reference | Continent |
|----------------------|---------------------------------------|--|--|-------------|--|--------------------------------|--------------------|-----------|
| Warming temperatures | Chinese Meteorological Administration | If 1980 variety was still grown, maize yield would stagnate or decrease; due to adoption of maize varieties with long growth period yield increased by 7-17% per decade. | China Agricultural Database | 1980-2009 | Heilongjiang Province, Northeast China | Single step attribution | (Meng et al. 2014) | Asia |
| Warming temperatures | Chinese Meteorological Administration | Changes in winter wheat phenology; observed dates of sowing, emergence, and beginning of winter dormancy were delayed by 1.2, 1.3, and 1.2 days per decade. Dates of regrowth after dormancy, anthesis, and maturity advanced 2.0, 3.7, and 3.1 days per decade. Growth duration, overwintering period, and vegetation phase shortened by 4.3, 3.1, and 5.0 days per decade. | Local agrometeorological experimental stations maintained by Chinese Meteorological Administration | 1981-2009 | Loess Plateau, Northwest China | Single step attribution | (He 2015) | Asia |

| | | | | | | | | |
|----------------------|-------------------------------------|--|--|-----------|------------------|--|---------------------|------|
| Warming temperatures | Central China Meteorological Agency | Advance in sowing and phenological stages advanced by 23-26 days | Agrometeorological experimental station Wulanwusu, China | 1981-2010 | Northwest China | Statistical relationships for cotton phenologies, seed cotton yields, and climate parameters using Pearson correlation analysis. | (Huang and Ji 2015) | Asia |
| Warming temperatures | China Meteorological Administration | Changes in temperature, precipitation and solar radiation in past three decades and increased wheat yield in northern China by 0.9-12.9%; reduced wheat yield in southern China by 1.2-10.2 %. | China Meteorological Administration | 1981-2009 | China | Correlations between annual yields with climate variables. Partial correlations with detrended yields and climate variables. | (Tao et al. 2014) | Asia |
| Warming temperatures | Pakistan Meteorological Department | Change in phenology of sunflowers. Sowing dates for spring sunflowers 3.4-9.3 days per decade earlier. Sowing dates for autumn sunflower delayed by 2.7-8.4 days per decade. | Punjab Agriculture Department | 1980-2016 | Punjab, Pakistan | Single step attribution | (Tariq et al. 2018) | Asia |

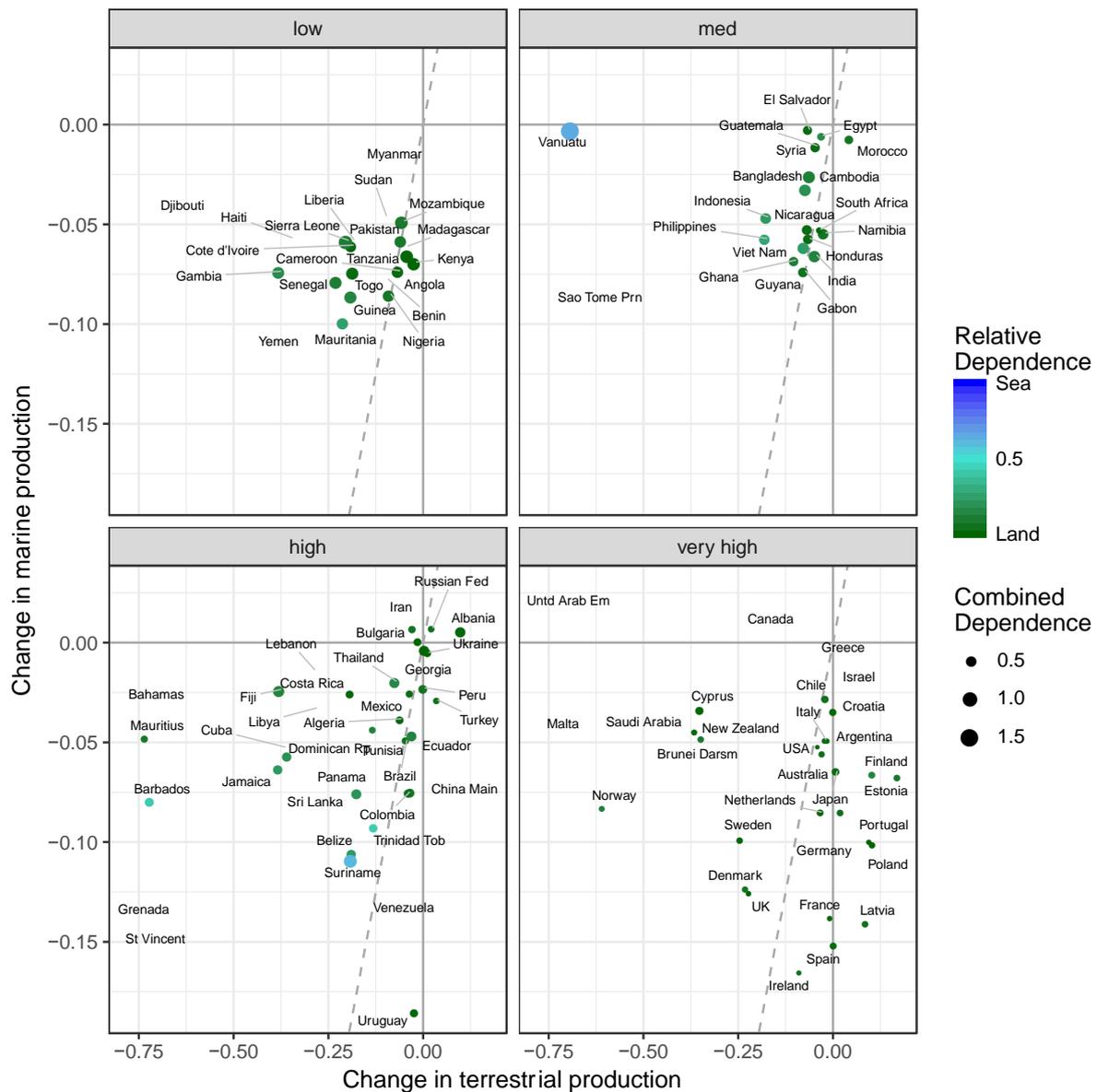
| | | | | | | | | |
|--|---------------------------------------|---|---|-----------|------------------|--|-----------------------|-----------|
| Warming temperatures | Pakistan Meteorological Department | Change in phenology in maize. Sowing dates for spring maize 3.5-5.5 days per decade earlier. Sowing dates for autumn maize 1.5-4.2 days per decade later. | Punjab Agriculture Department | 1980-2014 | Punjab, Pakistan | Single step attribution | (Abbas et al. 2017) | Asia |
| Increases in max and min temperatures | India Meteorological Department (IMD) | Reduced wheat yields by 5.2% . 1 degree C increase in maximum temperature lowers yields by 2.3% while same increase in minimum temperature lowers yields by 3.6%. | Indian Harvest Database Centre of Monitoring the Indian Economy (CMIE) and Directorate of Economics, Ministry of Agriculture. | 1981-2009 | India | Regression analysis between temperature and yield. | (Gupta et al. 2017) | Asia |
| Reduced rainfall and rising temperatures | Australian Bureau of Meteorology | Stagnated wheat yields. Declines in water-limited yield potential. | Agricultural Commodity Statistics | 1965-2015 | Australia | Single step attribution | (Hochman et al. 2017) | Australia |

| | | | | | | | | |
|--|--|--|--|-----------|----------------|---------------------------------|-------------------------|--------|
| Increases in temperature and drought | Czech Hydrometeorological Institute (CHMI), 268 climatological stations, and 774 rain gauge stations | Long-term impacts on fruiting vegetables (+4.9 to 12.2% per degree C) but decreases in stability of traditionally grown root vegetables in warmest areas of country. | Database of 12 field-grown vegetables at district level as reported by Czech Statistical Office. | 1961-2014 | Czech Republic | Associative pattern attribution | (Potopová et al. 2017) | Europe |
| Long-term temperature and precipitation trends | Precipitation: 1900-2008 Gridded Monthly Time Series Version 2.01. Available at: http://climate.geog.udel.edu/~climate/ . | Wheat and barley yields declined by 2.5% and 3.8%, and maize and sugar beet yields have increased due to temperature and precipitation changes. | EU Farm Accountancy Data Network (FADN) | 1989-2009 | Europe | Associative pattern attribution | (Moore and Lobell 2015) | Europe |

1 **Notes:** See Hegerl et al. (2010) for full definitions of attribution methods: Single Step: where a model(s) is run with and without a single variable of interest
2 (i.e., temperature) and results compared to observed changes within a system; Multi-Step: Through processes modelling and/or a statistical link, a change in
3 climate is linked to a variable of interest, and then that variable of interest is linked to an observed change; Associative Pattern: Involves the synthesis of
4 multiple observations – and demonstrates a pattern of strong association between these changes and changes in temperatures due to anthropogenic forcing.

5

1



2

3 **Figure SM5.1 Climate change impacts and adaptive capacity by continent across land and sea.**
 4 **Vulnerability of societies to climate change impacts in fisheries and agriculture under RCP6.0. Changes**
 5 **in marine fisheries (Tittensor 2017) and terrestrial crop production (Rosenzweig et al. 2014b) are**
 6 **expressed as $\log_{10}(\text{projected}/\text{baseline})$ production, where a value below zero indicates decreases and above**
 7 **are increases. Fisheries and agriculture dependency estimates calculated from employment, economy and**
 8 **food security. Circle size represents total dependency on both sectors and green to blue colour scale**
 9 **reflects the balance between land and sea with white indicative of equal dependence. The dependence**
 10 **indices were calculated using publicly available online data from FAO, the World Bank and a recent**
 11 **compilations of fisheries employment data (Teh and Sumaila 2013). Each panel a-d) represents the four**
 12 **Human Development Index (HDI) categories (low, medium, high and very high) and open diamonds**
 13 **indicate no data for agricultural and fisheries dependency. Modified from: Blanchard et al. 2017.**

14

15

16

17

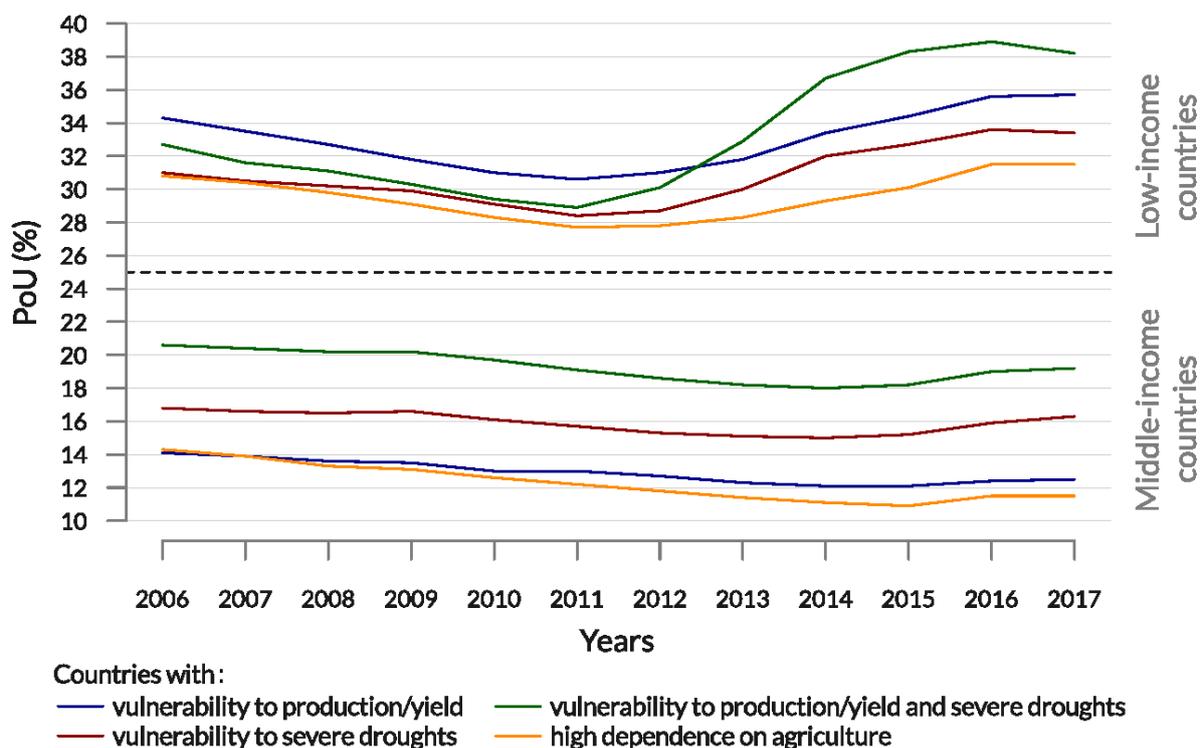
1

Table SM5.4 Models included in Hasegawa et al. (2018)

| Model | Reference |
|-----------|--|
| AIM/CGE | (Fujimori et al. 2012) |
| CAPRI | (Britz and Witzke 2014) |
| GCAM | (Kyle et al. 2011; Wise and Calvin 2011) |
| GLOBIOM | (Havlik et al. 2014) |
| IMAGE 3.0 | (Stehfest et al. 2014) |
| IMPACT 3 | (Robinson et al. 2015) |
| MAGNET | (Woltjer et al. 2014) |
| MAGPIE | (Lotze-Campen et al. 2008; Popp et al. 2014) |

2

3



4

5

6

Figure SM5.2 Undernourishment is higher when exposure to climate extremes is compounded by high levels of vulnerability in agriculture (FAO et al. 2018).

7

8

9

1 **Section SM5.5**

2 *Livestock mitigation strategies*

3 *Intensification of animal diets.* It is well established that appropriate diet regimes may contribute to
4 reduce the amount of GHG produced per unit of animal product (Gerber et al. 2013b), which, within
5 the appropriate implementation including governance, may lead to mitigation of absolute emissions.
6 This increased efficiency can be achieved through improved supplementation practices or through
7 land use management with practices like improved pasture management, including grazing rotation,
8 fertiliser applications, soil pH modification, development of fodder banks, improved pasture species,
9 use of legumes and other high protein feeds, the use of improved crop by-products and novel feeds
10 (i.e., black soldier fly meal, industrially produced microbial protein (Pikaar et al. 2018).

11 When done through increased feeding of grains, transition to improved diets shifts the contributions of
12 different GHG gases to the total emissions. This is due to the fact that the proportion of methane to
13 total emissions is reduced (due to lower roughage intake), while the proportion of emissions
14 associated with feed manufacture (energy and land use change) increases. Therefore, CO₂ emissions
15 from land use change increase while methane emissions per unit of output decrease (Gill et al. 2010).
16 As a consequence, the quantified benefits of a given strategy will also depend on the assumed GWP of
17 methane.

18 Of the available livestock GHG mitigation options, improved feeding systems are relatively easy to
19 implement at the farm level. A prerequisite for these options to work is that the livestock systems
20 need to be geared towards market-oriented production, as otherwise there is little incentive to improve
21 feeding systems. This in turn implies that costs and benefits to farmers are appropriate to incentivise
22 specific management changes and also assess the impact that market-orientation may have in some
23 societies, such as pastoralists (López-i-Gelats et al. 2016). Examples of where this option could be
24 applicable are smallholder dairy-crop mixed systems in Africa and Asia, dual-purpose and dairy
25 production in Latin America and beef cattle operations, where significant mitigation opportunities
26 exist. Other mitigation options include manipulation of rumen microflora, breeding for lower methane
27 production, and the use of feed additives (Hristov et al. 2013).

28 The largest GHG efficiency gaps are observed in livestock systems where the quality of the diet is the
29 poorest (i.e., grassland-based and some arid and humid mixed systems in the developing world). The
30 highest marginal gains of improving animal diets through simple feeding practices, both biologically
31 and economically, are in these systems (FAO, 2013; Herrero et al. 2013).

32 *Control of animal numbers, shifts in breeds, and improved management.* Increases in animal numbers
33 are one of the biggest factors contributing directly to GHG emissions (Tubiello, 2019). Regions with
34 intensive animal production, such as concentrated animal feeding operations (CAFOs), can control
35 animal numbers, conduct breeding programs for efficient animals, and improve feeding management.
36 In the developing world, many low-producing animals could be replaced by fewer but better-fed
37 cross-bred animals of a higher potential, with improved grazing management (i.e., attention to feed,
38 herbage availability, and allowances) playing an important role. In both developed and developing
39 countries these practices are able to reduce total emissions while maintaining or increasing the supply
40 of livestock products.

41 However, attention must be paid to synergies and trade-offs between livelihoods and specific
42 mitigation strategies, such as controlling animal numbers, recognising the multiple objectives that
43 livestock raising may contribute to within specific settings, especially in low-input systems.
44 Improvements in animal health can also significantly reduce emissions intensity by improved yields
45 and fertility per animal and reductions in mortality (ADAS 2015).

46 *Changes in livestock species.* Switching species to better suit particular environments is a strategy that
47 could yield higher productivity per animal for the resources available. At the same time, structural

1 changes in the livestock sector from beef to sheeps and goats, or mainly from ruminants to
2 monogastrics (e.g., from beef to pig or poultry production) could lead to reduced methane emissions
3 and higher efficiency gains. Assessment done using integrated assessment models (IAMs) have shown
4 that these practices could lead to reductions in land use change and its associated emissions (Havlik et
5 al. 2014; Frank et al. 2018).

6 *Managing nitrous oxide emissions from manure.* In the developing world, large amounts of nutrients
7 are lost due to poor manure management. In currently adopted feeding systems, large amounts of
8 nutrients and carbon are lost in connection with manure storage (e.g., Herrero et al. 2013). In many
9 places pig manure is not recycled; considered a waste, it is often discharged to water bodies or left to
10 accumulate unused. Yet these farming systems can be highly N and P limited. This practice creates
11 serious problems especially in urban and peri-urban systems by contributing to water and air
12 pollution. Research in intensive African ruminant livestock systems, for instance, has shown that up to
13 70% of the manure N can be lost within six months of excretion when manure is poorly managed
14 (Tittonell et al. 2009).

15 Options to manage emissions in the livestock sector are not easy to design because they require
16 systems thinking and awareness of key driving factors in different livestock systems. Reducing N
17 emissions starts with feeding livestock balanced diets so that excreta are not rich in labile N, which is
18 easily lost as ammonia and enters the N cascade (Bouwman et al. 2013). In intensive systems, mineral
19 N can be captured effectively using bedding material, which has been increasingly excluded from
20 livestock facilities to reduce operational costs.

21 Manure is increasingly handled as slurry in tanks or anaerobic lagoons, which may reduce direct
22 nitrous oxide emissions during storage but can increase methane and ammonia loss and also increase
23 the risk of emissions during land spreading (Velthof and Mosquera 2011). However, optimising land
24 spreading of manures (in terms of timing or placement) to maximise N and P replacement value can
25 minimise ammonia losses while also displacing mineral fertiliser (Bourdin et al. 2014).

26 In intensive systems, emissions of ammonia and nitrous oxide can be managed by spatially shifting
27 livestock pens or the facilities where they overnight. Other options in more-intensive grazing systems
28 may include nitrification inhibitors, stand-off pads, delayed manure spreading collected in milking
29 sheds, although the fate of the full applied N and its partitioning between direct and indirect emissions
30 as a result of the specific option chosen must be evaluated (e.g., Lam et al., 2017)

31

32 *Uncertainties in demand-side technical mitigation potential*

33 There are several unresolved issues regarding modelling and quantification of marginal emissions
34 identified in the literature. Diet shift studies often focus on beef production emission intensities,
35 although the cattle industry in many locations includes both meat and dairy production; these
36 activities may be integrated in different types of farming systems (Flysjö et al. 2012) with
37 significantly lower emission intensities (Gerber et al. 2013a; Flysjö et al. 2012). Links between
38 ruminant meat production, the dairy sector (primarily cows and goats), and wool production in sheep
39 are often overlooked in diet shift studies. FAOStat 2017 data indicate there are 278 million dairy cows
40 worldwide, which make significant contributions to meat production (304 million head slaughtered
41 per year) by providing calves (lactating cows must calve to produce milk) and dairy cows
42 (replacements by younger females).

43 Attributional LCA values are often applied to diet shifts studies, overlooking the feedback loop
44 (rebound effect) of demand on production system emission intensities. There are a few examples of
45 consequential analysis of diet shifts (Tukker et al. 2011) (de Oliveira Silva et al. 2016) (Zech and
46 Schneider 2019), reporting modest potential for mitigation (i.e., from 0-8%) but each of them

1 emphasise only one particular aspect of diet shifts. Further, the application of those models to
2 different regions of the world may require further development.

3 Current attributional LCA studies present inconsistencies related to the definition of system
4 boundaries, allocation of co-products (including dairy), method of attribution of land use change, and
5 pasture productivity effects on soil carbon stocks (Lynch 2019) (Yan et al. 2011; Dudley et al. 2014).
6 Major differences in the results are due to how land use change affects emissions and soil carbon
7 stocks, particularly when addressing developing countries where deforestation and intensification can
8 both take place at the same time. Deforestation-related emissions have been attributed to first land use
9 (Bustamante et al. 2012), the activities under a given amortization time (Persson et al. 2014), change
10 in total land covered by the activity (Gerber et al. 2013a), or the missed potential carbon sink, i.e., the
11 opportunity for natural vegetation recovery (Schmidinger and Stehfest 2012) (Schmidt et al. 2015).

12 Also, variation in soil carbon stocks is not considered in most studies, while a few account for
13 variations up to 0.3 m soil depth, and very rarely consider 1.0 m soil depth for estimating soil carbon
14 variation. Overlooking soil carbon at deeper soil layers largely contributes to underestimating the
15 environmental benefits of transition to more productive systems. Time considerations in soil carbon
16 stocks dynamics also vary among studies, with some applying a standard 20-year equilibrium time
17 instantaneously and others using dynamic (discrete or continuous) models.

18 The type of food replacement is another major source of uncertainty in calculating the impact of
19 dietary changes (Smetana et al. 2015). Nutritional replacement with animal-based protein candidates
20 such as chicken, eggs, pork, fish, and insects is likely to vary widely in different geographical
21 contexts. While chicken and soybean are currently dominating international trade of protein sources
22 (FAOStat), legumes, pulses, seaweed, and yeast-derived foods are being tested as ingredients by the
23 food industry.

24 In regard to food quality, reducing meat consumption may lower the iron and zinc nutritional status of
25 certain vulnerable groups. For example, in Europe 22% of preschool children, 25% of pregnant
26 women, and 19% of nonpregnant women already have anemia (WHO, 2008). Reductions in red meat
27 consumption also may have food safety implications. Substituting meat with poultry or seafood might
28 increase foodborne illnesses, whereas replacement with pulses and vegetables would reduce them
29 (Lake et al., 2012).

30 GHG emissions associated with food preparation and food waste are usually unaccounted for in diet
31 shift studies with rare exceptions (Corrado et al. 2019). Dietary supplements (vitamin, minerals and
32 amino acids) are highly recommended for low-meat diets, but they are not considered in GHG
33 mitigation studies of diet shifts, mostly because of lack of LCA data for supplements (Corrado et al.
34 2019).

35 The varying proportions of CO₂, CH₄, and N₂O contributions to ruminant-related emissions, with a
36 high proportion of the short-lived methane, make interpretation sensitive to the global warming
37 metrics adopted (Reisinger and Clark 2018) (Lynch 2019). As more intensive systems or other diet
38 alternatives would alter the relative contributions to food of these gases, the choice of metric often
39 changes the ranking of mitigation options (Lynch and Pierrehumbert 2019)(Garnett, 2011). Most
40 projections related to diet shifts do not account for the potential of methane inhibitors, non-symbiotic
41 nitrogen fixation, advances in livestock and forage genetics, and other emerging technologies in the
42 livestock sector, some of which are close to market launch (Jayanegara et al. 2018).

43 In a systems view, dairy and wool production can be affected if reductions in ruminant meat demand
44 take place. While beef production systems are often characterised by low energy and protein
45 efficiency, milk production is as efficient energetically as egg production and second after eggs in
46 protein conversion efficiency among animal-based proteins (Eshel et al. 2016).

1 In summary, systems level analyses revealed wide variation in mitigation estimates of diet shifts, in
2 part due to differing accounting for the main interactions. There is *robust evidence* that diet shifts can
3 mitigate GHG emissions but *low agreement* on how much could be achieved and what would be the
4 effectiveness of interventions to promote diet shifts. In high-income industrialised countries, there is
5 scope for reducing consumption of livestock produce with tangible environmental benefits; in
6 developing countries, high meat-based diets are less prevalent and scope for reductions may be more
7 limited, but there are options for encouraging nutrition transitions towards healthy diets.

8

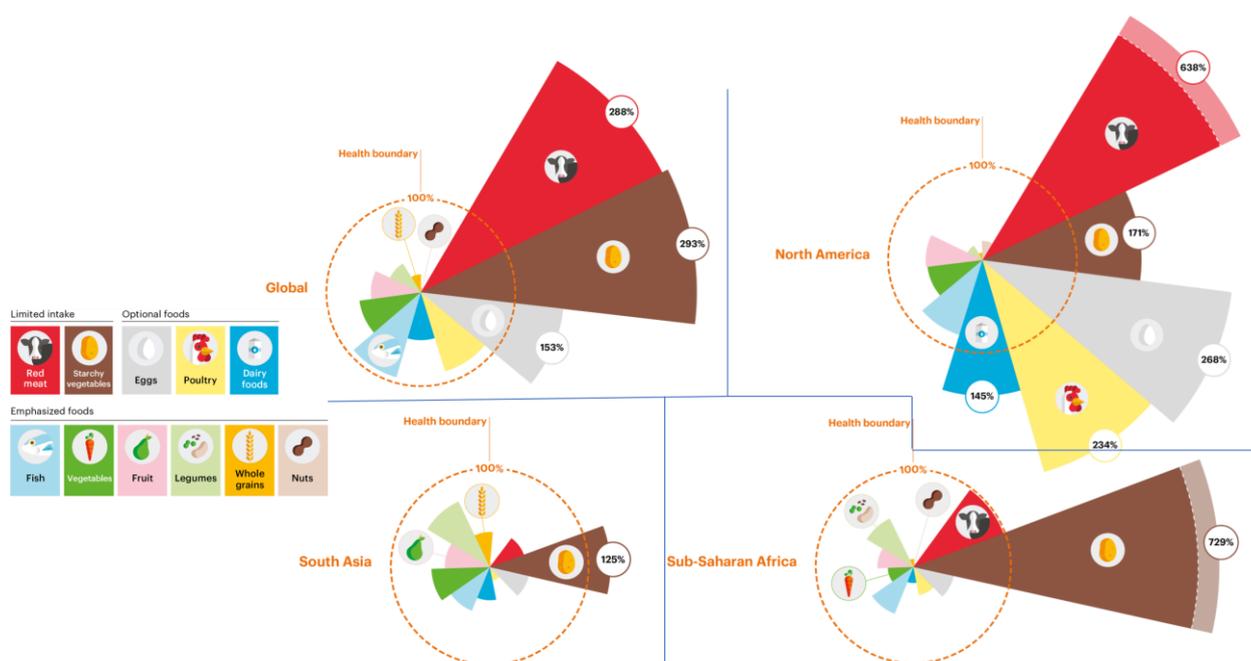
9

1 Section SM5.6

2 *Global meat consumption*

3 The issue of global meat consumption as a driver of GHG emission, can be weighed against the
 4 requirements of healthy diet. Healthy and sustainable diets are high in coarse grains, pulses, fruits and
 5 vegetables, and nuts and seeds; low in energy-intensive animal-sourced and discretionary foods (such
 6 as sugary beverages and fats); and have a carbohydrate threshold. Based on the potential impact of
 7 suboptimal diets on non-communicable diseases (NCD) mortality and morbidity, the World Health
 8 Organization (WHO) and the EAT-LANCET report (Willett et al. 2019) highlighted the need for
 9 improving diets across nations and made recommendations on how to balance nutrition to prevent
 10 malnutrition. The source of protein is not limited to meat; it is found in fish, vegetable and insects.
 11 The range of options in balancing protein sources runs primarily into cultural resistance, food habits,
 12 economic conditions and the social and economic factors influencing how the food system affects
 13 climate and land.

14 Most recent analyses, like the EAT-LANCET (Willett et al. 2019) work, show that reductions in
 15 consumption, especially of red meat, apply to over-consumers, while scope remains for growth in
 16 consumption in Low- and Middle-Income Countries (LMICs).



17

18 **Figure SM5.3 The “diet gap” between current dietary patterns and intakes of food in the planetary health**
 19 **diet (Willett et al. 2019).**

20 From the climate and land perspectives, there is a difference between red meat production and other
 21 meat production (Willett et al. 2019). The impacts of meat production will depend on resource use
 22 intensity to produce meat calories, the land and climate footprints of the processing and supply chains,
 23 and the scale of the production systems (i.e., livestock on crop by-products vs. pasture vs. intensive
 24 grain-fed) (Willett et al. 2019). Hence, the question is not about eating less meat for everyone, but to
 25 adopt sustainable supply and consumption practices across a broad range of food systems.

26 The biggest challenge to achieve changes in meat consumption is on how to start a transition that has
 27 increasing diversity of food sources with lower land and water requirements and GHG emissions. This
 28 could be a gradual transition that recognises the need for just transitions for people whose livelihoods
 29 depend on (red) meat production. In this regard, all parts of the food system, including production,
 30 trade, and consumption, play important roles.

Subject to Copy-editing

1

2 **Section SM5.7**

3 ***Governance***

4 Governance of climate change and governance of food systems have been developed independently of
5 each other. This section highlights the main characteristics of food and climate governance and
6 assesses what options may exist for establishing arrangements that link the two. See Chapter 7 for
7 important characteristics of governance and institutions; here we describe those relevant for
8 enhancing the interactions between climate change and food systems.

9 In the governance of climate change, Huitema et al. (2016) highlighted differences between mitigation
10 and adaptation. Mitigation often requires global agreements and national policies while adaptation
11 requires local and regional considerations. However, in the case of food systems this difference does
12 not apply, because mitigation measures also require local actions (e.g., at the farm level), while
13 adaptation actions may also require measures at global and national levels (such as emergency food
14 aid for climate disasters and food safety nets).

15 Governance of food systems holds particular challenges because it is only recently that a systems
16 approach has been embraced by policy-makers. (Rivera-Ferre et al. 2013) proposed principles for
17 food systems management considering them as complex socioecological systems (SES) including:
18 learning, flexibility, adaptation, participation, diversity enhancement, and precaution. These principles
19 are part of the framework of adaptive governance (see Chapter 7). Termeer et al. (2018) developed a
20 diagnostic framework with five principles to assess governance options appropriate to food systems:
21 1) system-based problem framing; 2) connectivity across boundaries to span siloed governance
22 structures and include non-state actors; 3) adaptability to flexibly respond to inherent uncertainties
23 and volatility; 4) inclusiveness to facilitate support and legitimacy; and 5) transformative capacity to
24 overcome path dependencies and create conditions to foster structural change.

25 Both the food and climate systems require integrated governance and institutions (*high confidence*).
26 These need to span government levels and actors across a wide range of sectors including agriculture,
27 environment, economic development, health, education, and welfare (Misselhorn et al. 2012). For
28 climate and food system management, the creation of government entities or ministerial units
29 responsible for coordinating among these ministries (horizontal coordination) and for cutting across
30 different administrative levels (vertical coordination) have been proposed (Orr et al. 2017).

31 However, integration is not easy. Termeer et al. (2018) analysed three South African governance
32 arrangements that explicitly aim for a holistic system-based approach. They found that they were not
33 delivering the expected outcomes due to reversion to technical one-dimensional problem framing.
34 Issues included dominance of single departments, limited attention to monitoring and flexible
35 responses, and exclusion of those most affected by food insecurity. Newell et al. (2018) analysed the
36 governance process of climate smart agriculture (CSA) from global to local scales for Kenya and
37 found a triple disconnect between global, national, and local scales. Different levels of authority and
38 actors imposed their own framing of CSA, and how to implement it.. As a result of the competition
39 among different actors, siloed policy practices were reproduced.

40 Food systems governance must also include governance of the resources needed to produce food,
41 which vary from land tenure (see chapter 7) and seed sovereignty (see Chapter 6), to other resources
42 such as soil fertility. Montanarella and Vargas (2012) proposed a supranational structure to guarantee
43 soil conservation on all continents, such as the Global Soil Partnership. This can also apply for the
44 governance of food and climate systems.

45 Polycentric and multiscale governance structures have been proposed for coping with climate change
46 to address both mitigation and adaptation (Ostrom 2010), and were suggested by Rivera-Ferre et al.

1 (2013) for food systems. A polycentric approach provides more opportunities for experimentation and
2 learning across levels (Cole 2015), entails many policy experiments from which policymakers at
3 various levels of governance can learn (Ostrom 2010), and contributes to building trust among
4 stakeholders (e.g., nation states, public and private sectors, civil society). Polycentric approaches
5 have been suggested for the Sustainable Development Goals (SDGs) (Monkelbaan 2019).

6 Another governance option suggested for the SDGs (Monkelbaan 2019) are already implemented in
7 global atmospheric and marine agreements (e.g., the Montreal protocol (De Búrca et al. 2014; Armeni
8 2015) is global experimentalist governance). Global experimentalist governance is an institutionalised
9 process of participatory and multilevel collective problem-solving, in which the problems (and the
10 means of addressing them) are framed in an open-ended way, and subjected to periodic revision by
11 peer review in the light of locally generated knowledge (De Búrca et al. 2014), This favours learning,
12 participation and cooperation (Armeni 2015). This form of governance can establish processes that
13 enable unimagined alternatives.

15 ***Institutions***

16 As Candel (2014) highlighted, based on a systematic review of food security governance focused on
17 hunger, global governance of food security is lacking because there is no institution with a mandate to
18 address concerns across sectors and levels. No international organisation deals with food security in a
19 holistic and inclusive manner. This results in overlapping (often conflicting) norms, rules and
20 negotiations that generate a “regime complex” (Margulis 2013), particularly in regard to agriculture
21 and food, international trade and human rights (e.g. UN Committee of World Food Security (CFS),
22 WTO, G8, G20). In climate change governance there are also multiple overlapping institutions with
23 often-conflicting rules and actors (Keohane and Victor 2011).

24 New multi-stakeholder governance arrangements are emerging, such as the Global Agenda for
25 Sustainable Livestock (Breeman et al. 2015) and the CFS (Duncan 2015). Also relevant in food
26 systems and climate change governance is that food security governance is spread across domains,
27 sectors and spatial scales (global, regional, national, local, community, household, or individual) with
28 a lack of coherency and coordination across multiple scales (*high confidence*). Thus, a major
29 challenge is to coordinate all these domains, sectors and scales.

30 It is important to consider the variety of actors involved in food security governance at all levels
31 (international bodies, civil society organisations (CSOs), nation states, public sector groups, and
32 private sector entities), with different agendas and values. But new in this regard is the participation of
33 CSOs that can provide the policy-making process with bottom-up knowledge to identify food
34 insecurity issues and locally relevant responses. CSOs can also contribute to multi-sector and multi-
35 scalar approaches by bridging government agencies and levels (Candel 2014). Thus, to facilitate
36 coordination and coherence, new adaptive governance enables interactions across multiple levels and
37 scales (Pereira and Ruysenaar 2012) and the use of “boundary organisations” (Candel 2014). To
38 address different narratives regarding food security (Rivera-Ferre 2012; Lang and Barling 2012), a
39 first step is to agree on basic principles and values (Margulis 2013).

40 In this regard, an opportunity to address food systems governance challenges arises within the UN
41 Committee on World Food Security (CFS), where diverse actors, voices and narratives are integrated
42 in the global food security governance. As a point of departure, the CFS could provide the platform to
43 develop global experimentalist governance in food systems (Duncan 2015; Duncan and Barling 2012)
44 providing a combination of bottom-up and top-down initiatives (Lambek 2019). However, the
45 existence of overlapping structures with different focuses on food security and power may hinder the
46 potential of this institution. (Margulis and Duncan 2016).

1 Mainstreaming of collaborative and more inclusive modes of governance, such as those displayed at
2 the CFS, are needed to effectively address the impacts of a changing planet on food systems
3 (Barling and Duncan 2015) and improve the balance of sustainable production and food consumption.
4 Despite improvements in global food security, food systems and climate governance, the main focus
5 is still on food security as undernutrition. New challenges will arise from the increasing evidence of
6 the burden of obesity, for which other institutions, focused on nutrition, will be needed. The new
7 Global Strategy Framework for Food Security and Nutrition (Committee on World Food Security
8 2017) of the CFS provides a new overarching framework for food security and nutrition strategies,
9 policies and actions that includes environmental concerns within a food system approach and a broad
10 vision of food and nutrition security. This framework fits within the “governance through goals”
11 provided by the SDGs (Biermann et al. 2017).

12 Both in climate change and food systems, the sub-national governance at the level of cities and
13 communities is also becoming relevant in terms of responses (*high evidence, high agreement*). From a
14 climate change perspective (see Chapter 7 for more examples) transnational municipal networks,
15 particularly transnational municipal climate networks, have played a key role in climate change
16 mitigation and have potential to facilitate adaptation (Fünfgeld 2015; Busch et al. 2018; Rosenzweig
17 et al. 2018). Efficient food systems require subnational governments to include food policy councils
18 (Feenstra 2002; Schiff 2008) and cities networks to address food systems challenges (e.g., Sustainable
19 Food Cities in the UK or Agroecological Cities in Spain). Transition Towns are engaged in common
20 principles towards sustainable development, including food systems transformation for food security
21 (Sage 2014), health and well-being (Richardson et al. 2012), and climate change (Taylor Aiken 2015).

22

23 ***Scope for expanded policies***

24 The interaction of production-based support through agricultural policy, coupled with agricultural
25 research investment and the development of frameworks to liberalise trade has led to a range of
26 consequences for global and local food systems. Together, these policies have shaped the food system
27 and incentivised global intensification of agriculture, and significant gains in global production.
28 However, jointly they have also incentivised a concentration on a small number of energy-dense
29 commodity crops grown at large scales (*high confidence*) (just eight crops supply 75% of the world’s
30 consumed calories (West et al. 2014)). The production of these commodity crops underpin global
31 dietary transitions, leading to dietary homogenisation (based primarily on starchy grains/tubers,
32 vegetable oil, sugar and livestock produce) (Khoury et al. 2014).

33 Global intensification of agriculture, as well as increasing the supply of affordable calories, has
34 impacted soil, water, air quality and biodiversity in major and negative ways (Dalin et al. 2017;
35 Tamea et al. 2016; Newbold et al. 2015; García-Ruiz et al. 2015; Amundson et al. 2015; Paulot and
36 Jacob 2014). Importantly in the context of this report, a narrow focus on productivity has led to a food
37 system that emits a large proportion of GHGs (Section 5.4), is fragile in the face of climate shocks
38 (Section 5.3) and from which food is used inefficiently (through waste and over-consumption, Section
39 5.5.2.5). Mitigation of climate change, as well as adaptation, can then arise from a transformation of
40 the food system to one that provides nutrition and health (Willett et al. 2019; Springmann et al.
41 2018b,a; Godfray et al. 2018; Ramankutty et al. 2018; Chaudhary et al. 2018). There is therefore
42 *medium confidence*, that continued focus on the past drivers of the food system will be detrimental for
43 climate change and food security.

44 Addressing this challenge requires action across the food system to enhance synergies and co-benefits
45 and minimise trade-offs among multiple objectives of food security, adaptation and mitigation
46 (Sapkota et al. 2017; Palm et al. 2010; Jat et al. 2016; Sapkota et al. 2015) (Section 5.6), as well as
47 broader environmental goods exemplified by the SDG framework such as water, air-quality, soil
48 health and biodiversity (Obersteiner et al. 2016; Pradhan et al. 2017). In short, this requires greater

1 policy alignment and coherence between traditionally separate policy domains to recognise the
2 systemic nature of the problem. For example, aligning the policy goals of sustainable land
3 management for the purposes of managing both food security and biodiversity (Meyfroidt 2017;
4 Wittman et al. 2017), or public health and agricultural policies (Thow et al. 2018) that can drive
5 mitigation, as well as the enabling conditions of land rights, tenure and ownership. Significant co-
6 benefits can arise from integrated food systems policies, as well as integrated approaches to
7 generating evidence to underpin coherent policy, exemplified, for example, by the EU’s integrated
8 research and innovation strategy “Food2030” that aligns agriculture, environment, nutrition and
9 research policy (European Commission 2018).

10

1 **References**

- 2 Abbas, G., and Coauthors, 2017: Quantification the impacts of climate change and crop management
3 on phenology of maize-based cropping system in Punjab, Pakistan. *Agric. For. Meteorol.*, **247**,
4 42–55, doi:10.1016/j.agrformet.2017.07.012.
5 <https://linkinghub.elsevier.com/retrieve/pii/S0168192317302289> (Accessed November 4, 2018).
- 6 Aberman, N.-L., and C. Tirado, 2014: Impacts of Climate Change on Food Utilization. *Global*
7 *Environmental Change*, B. Freedman, Ed., Springer, Dordrecht, 717–724
8 <http://link.springer.com/10.1007/978-94-007-5784-4>.
- 9 Abid, M., U. A. Schneider, and J. Scheffran, 2016: Adaptation to climate change and its impacts on
10 food productivity and crop income: Perspectives of farmers in rural Pakistan. *J. Rural Stud.*, **47**,
11 254–266, doi:10.1016/J.JRURSTUD.2016.08.005.
12 <https://www.sciencedirect.com/science/article/pii/S0743016716302911> (Accessed October 1,
13 2018).
- 14 Abiona, B. G., E. O. Fakoya, and J. Esun, 2016: The Impacts of Climate Change on the Livelihood of
15 Arable Crop Farmers in Southwest, Nigeria. Springer, Cham, 289–296
16 http://link.springer.com/10.1007/978-3-319-25814-0_20 (Accessed November 4, 2018).
- 17 ADAS, 2015: *Study to Model the Impact of Controlling Endemic Cattle Diseases and Conditions on*
18 *National Cattle Productivity, Agricultural Performance and Greenhouse Gas Emissions*. 210 pp.
19 http://sciencesearch.defra.gov.uk/Document.aspx?Document=13320_AC0120Finalreport.pdf.
- 20 Amundson, R., A. A. Berhe, J. W. Hopmans, C. Olson, A. E. Sztein, and D. L. Sparks, 2015: Soil and
21 human security in the 21st century. *Science (80-.)*, **348**, doi:10.1126/science.1261071.
- 22 Armeni, C., 2015: GLOBAL EXPERIMENTALIST GOVERNANCE, INTERNATIONAL LAW
23 AND CLIMATE CHANGE TECHNOLOGIES. *Int. Comp. Law Q.*, **64**, 875–904,
24 doi:10.1017/S0020589315000408.
25 http://www.journals.cambridge.org/abstract_S0020589315000408 (Accessed March 2, 2019).
- 26 Asseng, S., and Coauthors, 2013: Uncertainty in simulating wheat yields under climate change. *Nat.*
27 *Clim. Chang.*, **3**, 827–832, doi:10.1038/nclimate1916.
28 <http://www.nature.com/articles/nclimate1916> (Accessed April 16, 2019).
- 29 —, and Coauthors, 2015: Rising temperatures reduce global wheat production. *Nat. Clim. Chang.*,
30 **5**, 143–147, doi:10.1038/nclimate2470. <http://www.nature.com/doifinder/10.1038/nclimate2470>.
- 31 Bahrami, H., L. J. De Kok, R. Armstrong, G. J. Fitzgerald, M. Bourgault, S. Henty, M. Tausz, and S.
32 Tausz-posch, 2017: The proportion of nitrate in leaf nitrogen , but not changes in root growth ,
33 are associated with decreased grain protein in wheat under elevated [CO₂]. *J. Plant Physiol.*,
34 **216**, 44–51, doi:10.1016/j.jplph.2017.05.011.
- 35 Barling, D., and J. Duncan, 2015: The dynamics of the contemporary governance of the world’s food
36 supply and the challenges of policy redirection. *Food Secur.*, **7**, 415–424, doi:10.1007/s12571-
37 015-0429-x. <http://link.springer.com/10.1007/s12571-015-0429-x> (Accessed March 2, 2019).
- 38 Bassu, S., and Coauthors, 2014: How do various maize crop models vary in their responses to climate
39 change factors? *Glob. Chang. Biol.*, **20**, 2301–2320, doi:10.1111/gcb.12520.
- 40 Bebbler, D. P., T. Holmes, and S. J. Gurr, 2014: The global spread of crop pests and pathogens. *Glob.*
41 *Ecol. Biogeogr.*, **23**, 1398–1407, doi:10.1111/geb.12214.
- 42 Biermann, F., N. Kanie, and R. E. Kim, 2017: Global governance by goal-setting: the novel approach
43 of the UN Sustainable Development Goals. *Curr. Opin. Environ. Sustain.*, **26–27**, 26–31,
44 doi:10.1016/J.COSUST.2017.01.010.
45 <https://www.sciencedirect.com/science/article/pii/S1877343517300209> (Accessed February 27,
46 2019).
- 47 Bourdin, F., R. Sakrabani, M. G. Kibblewhite, and G. J. Lanigan, 2014: Effect of slurry dry matter

- 1 content, application technique and timing on emissions of ammonia and greenhouse gas from
2 cattle slurry applied to grassland soils in Ireland. *Agric. Ecosyst. Environ.*, **188**, 122–133,
3 doi:10.1016/j.agee.2014.02.025. <http://linkinghub.elsevier.com/retrieve/pii/S0167880914001066>
4 (Accessed May 30, 2018).
- 5 Bouwman, L., and Coauthors, 2013: Exploring global changes in nitrogen and phosphorus cycles in
6 agriculture induced by livestock production over. *Proc Natl Acad Sci*, **110**, 20882–20887,
7 doi:10.1073/pnas.1206191109. www.pnas.org/cgi/doi/10.1073/pnas.1206191109.
- 8 Breeman, G., J. Dijkman, and C. Termeer, 2015: Enhancing food security through a multi-stakeholder
9 process: the global agenda for sustainable livestock. *Food Secur.*, **7**, 425–435,
10 doi:10.1007/s12571-015-0430-4.
- 11 Britz, W., and P. Witzke, 2014: *CAPRI model documentation*.
- 12 De Búrca, G., R. O. Keohane, and C. Sabel, 2014: Global Experimentalist Governance. *Br. J. Polit.*
13 *Sci.*, **44**, 477–486, doi:10.1017/S0007123414000076.
14 http://www.journals.cambridge.org/abstract_S0007123414000076 (Accessed February 26,
15 2019).
- 16 Busch, H., L. Bendlin, and P. Fenton, 2018: Shaping local response – The influence of transnational
17 municipal climate networks on urban climate governance. *Urban Clim.*, **24**, 221–230,
18 doi:10.1016/j.uclim.2018.03.004.
- 19 Bustamante, M. M. C., and Coauthors, 2012: Estimating greenhouse gas emissions from cattle raising
20 in Brazil. *Clim. Change*, **115**, doi:10.1007/s10584-012-0443-3.
- 21 Candel, J. J. L., 2014: Food security governance: a systematic literature review. *Food Secur.*, **6**, 585–
22 601, doi:10.1007/s12571-014-0364-2. <http://link.springer.com/10.1007/s12571-014-0364-2>
23 (Accessed February 26, 2019).
- 24 Chaudhary, A., D. Gustafson, and A. Mathys, 2018: Multi-indicator sustainability assessment of
25 global food systems. *Nat. Commun.*, **9**, 848, doi:10.1038/s41467-018-03308-7.
26 <http://www.nature.com/articles/s41467-018-03308-7> (Accessed March 4, 2019).
- 27 Chen, Y., and Coauthors, 2018: Great uncertainties in modeling grazing impact on carbon
28 sequestration: a multi-model inter-comparison in temperate Eurasian Steppe. *Environ. Res. Lett.*,
29 **13**, 075005, doi:10.1088/1748-9326/aacc75. <http://stacks.iop.org/1748-9326/13/i=7/a=075005?key=crossref.8c1ea2666d647d78ac753b12554f5bb7>.
- 31 Cole, D. H., 2015: Advantages of a polycentric approach to climate change policy. *Nat. Clim. Chang.*,
32 **5**, 114–118, doi:10.1038/nclimate2490. <http://www.nature.com/articles/nclimate2490> (Accessed
33 February 26, 2019).
- 34 Committee on World Food Security, 2017: *Global Strategic Framework for Food Security and*
35 *Nutrition*. <http://www.fao.org/cfs/home/activities/gsf/en/> (Accessed March 2, 2019).
- 36 Corrado, S., G. Luzzani, M. Trevisan, and L. Lamastra, 2019: Contribution of different life cycle
37 stages to the greenhouse gas emissions associated with three balanced dietary patterns. *Sci. Total*
38 *Environ.*, **660**, 622–630, doi:10.1016/J.SCITOTENV.2018.12.267.
- 39 Cramer, W., G. W. Yohe, M. Auffhammer, C. Huggel, U. Molau, M. A. F. da S. Dias, and R.
40 Leemans, 2014: Detection and attribution of observed impacts. *Climate Change 2014:*
41 *Impacts, Adaptation, and Vulnerability*, Cambridge University Press, 979–1038.
- 42 Dalin, C., Y. Wada, T. Kastner, and M. J. Puma, 2017: Groundwater depletion embedded in
43 international food trade. *Nature*, **543**, 700–704, doi:10.1038/nature21403.
44 <http://dx.doi.org/10.1038/nature21403>.
- 45 Dudley, Q. M., A. J. Liska, A. K. Watson, and G. E. Erickson, 2014: Uncertainties in life cycle
46 greenhouse gas emissions from U.S. beef cattle. *J. Clean. Prod.*, **75**, 31–39,
47 doi:10.1016/J.JCLEPRO.2014.03.087.

- 1 Duncan, J., 2015: *Global Food Security Governance*. Routledge,
2 <https://www.taylorfrancis.com/books/9781317623205> (Accessed March 2, 2019).
- 3 Duncan, J., and D. Barling, 2012: Renewal through Participation in Global Food Security
4 Governance: Implementing the International Food Security and Nutrition Civil Society
5 Mechanism to the Committee on World Food Security. <http://openaccess.city.ac.uk/2577/>
6 (Accessed February 28, 2019).
- 7 EPA, 2012: Global Anthropogenic Non-CO2 Greenhouse Gas Emissions: 1990-2030. *Off. Atmos.*
8 *Programs Clim. Chang. Div. U.S. Environ. Prot. Agency*, doi:EPA 430-R-12-006.
- 9 Eshel, G., A. Shepon, E. Noor, and R. Milo, 2016: Environmentally Optimal, Nutritionally Aware
10 Beef Replacement Plant-Based Diets. *Environ. Sci. Technol.*, **50**, 8164–8168,
11 doi:10.1021/acs.est.6b01006.
- 12 European Commission, 2018: *Recipe for change: An agenda for a climate-smart and sustainable food*
13 *system for a healthy Europe*. Brussels, 1-140 pp. [https://publications.europa.eu/en/publication-](https://publications.europa.eu/en/publication-detail/-/publication/d0c725de-6f7c-11e8-9483-01aa75ed71a1/language-en)
14 [detail/-/publication/d0c725de-6f7c-11e8-9483-01aa75ed71a1/language-en](https://publications.europa.eu/en/publication-detail/-/publication/d0c725de-6f7c-11e8-9483-01aa75ed71a1/language-en) (Accessed April 17,
15 2019).
- 16 FAO, 2007: *The State of Food and Agriculture 2007*. Rome, Italy, 222p pp.
- 17 —, 2016: *The State of Food and Agriculture. Climate Change, Agriculture, and Food Security*.
18 Rome, 194 pp. <http://www.fao.org/3/a-i6030e.pdf> (Accessed October 1, 2018).
- 19 —, 2018: *The future of food and agriculture: Trends and challenges*. Rome, 180 pp.
20 <http://www.fao.org/3/a-i6583e.pdf> (Accessed May 18, 2018).
- 21 —, IFAD, UNICEF, WFP, and WHO, 2018: *The State of Food Security and Nutrition in the World*
22 *2018. Building climate resilience for food security and nutrition*. Rome,
23 www.fao.org/publications (Accessed October 1, 2018).
- 24 Feenstra, G., 2002: Creating space for sustainable food systems: Lessons from the field. *Agric.*
25 *Human Values*, **19**, 99–106, doi:10.1023/A:1016095421310.
26 <http://link.springer.com/10.1023/A:1016095421310> (Accessed February 26, 2019).
- 27 Flysjö, A., C. Cederberg, M. Henriksson, and S. Ledgard, 2012: The interaction between milk and
28 beef production and emissions from land use change – critical considerations in life cycle
29 assessment and carbon footprint studies of milk. *J. Clean. Prod.*, **28**, 134–142,
30 doi:10.1016/J.JCLEPRO.2011.11.046.
- 31 Frank, S., and Coauthors, 2018: Structural change as a key component for agricultural non-CO2
32 mitigation efforts. *Nat. Commun.*, **9**, 1060, doi:10.1038/s41467-018-03489-1.
33 <http://www.nature.com/articles/s41467-018-03489-1>.
- 34 Fratkin, E., E. A. Roth, and M. A. Nathan, 2004: Pastoral Sedentarization and Its Effects on
35 Children's Diet, Health, and Growth Among Rendille of Northern Kenya. *Hum. Ecol.*, **32**, 531–
36 559, doi:10.1007/s10745-004-6096-8. <http://link.springer.com/10.1007/s10745-004-6096-8>
37 (Accessed March 24, 2019).
- 38 Fujimori, S., T. Masui, and M. Yuzuru, 2012: *AIM/CGE [basic] manual*. Tokuba,.
- 39 Fünfgeld, H., 2015: Facilitating local climate change adaptation through transnational municipal
40 networks. *Curr. Opin. Environ. Sustain.*, **12**, 67–73, doi:10.1016/J.COSUST.2014.10.011.
41 <https://www.sciencedirect.com/science/article/pii/S1877343514000839> (Accessed February 26,
42 2019).
- 43 García-Ruiz, J. M., S. Beguería, E. Nadal-Romero, J. C. González-Hidalgo, N. Lana-Renault, and Y.
44 Sanjuán, 2015: A meta-analysis of soil erosion rates across the world. *Geomorphology*, **239**,
45 160–173, doi:http://dx.doi.org/10.1016/j.geomorph.2015.03.008.
46 <http://www.sciencedirect.com/science/article/pii/S0169555X1500149X>.
- 47 Gerber, P. J., Steinfeld, H., Henderson, B., Mottet, A., Opio, C., and J. Dijkman, 2013a: *Animal*

- 1 *Production and Health Division A global life cycle assessment Greenhouse gas emissions from*
2 *ruminant supply chains.*
- 3 Gerber, P. J., H. Steinfeld, B. Henderson, and others, 2013b: *Tackling climate change through*
4 *livestock: a global assessment of emissions and mitigation opportunities.* Food and Agriculture
5 Organization of the United Nations, Rome,.
- 6 Gill, M., P. Smith, and J. M. Wilkinson, 2010: Mitigating climate change: the role of domestic
7 livestock. *Animal*, **4**, 323–333, doi:10.1017/S1751731109004662.
- 8 Godfray, H. C. J., and Coauthors, 2018: Meat consumption, health, and the environment. *Science* (80-
9), **361**, eaam5324, doi:10.1126/science.aam5324.
10 <http://www.ncbi.nlm.nih.gov/pubmed/30026199> (Accessed November 4, 2018).
- 11 Gupta, R., E. Somanathan, and S. Dey, 2017: Global warming and local air pollution have reduced
12 wheat yields in India. *Clim. Change*, **140**, 593–604, doi:10.1007/s10584-016-1878-8.
13 <http://link.springer.com/10.1007/s10584-016-1878-8> (Accessed May 17, 2018).
- 14 Harvey, C. A., Z. L. Rakotobe, N. S. Rao, R. Dave, H. Razafimahatratra, R. H. Rabarijohn, H.
15 Rajaofara, and J. L. Mackinnon, 2014: Extreme vulnerability of smallholder farmers to
16 agricultural risks and climate change in Madagascar. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*,
17 **369**, 20130089, doi:10.1098/rstb.2013.0089. <http://www.ncbi.nlm.nih.gov/pubmed/24535397>
18 (Accessed October 1, 2018).
- 19 Hasegawa, T., and Coauthors, 2018: Risk of increased food insecurity under stringent global climate
20 change mitigation policy. *Nat. Clim. Chang.*, **8**, 699–703, doi:10.1038/s41558-018-0230-x.
21 <http://www.nature.com/articles/s41558-018-0230-x> (Accessed August 16, 2018).
- 22 Havlik, P., and Coauthors, 2014: Climate change mitigation through livestock system transitions.
23 *Proc. Natl. Acad. Sci.*, **111**, 3709–3714, doi:10.1073/pnas.1308044111.
24 <http://www.pnas.org/cgi/doi/10.1073/pnas.1308044111>.
- 25 He, J., 2015: Chinese public policy on fisheries subsidies: Reconciling trade, environmental and food
26 security stakes. *Mar. Policy*, **56**, 106–116, doi:10.1016/j.marpol.2014.12.021.
27 <http://dx.doi.org/10.1016/j.marpol.2014.12.021>.
- 28 Hegerl, G. C., O. Hoegh-Guldberg, G. Casassa, M. P. Hoerling, R. S. Kovats, C. Parmesan, D. W.
29 Pierce, and P. A. Stott, 2010: Good practice guidance paper on detection and attribution related
30 to anthropogenic climate change. *Meeting Report of the Intergovernmental Panel on Climate*
31 *Change Expert Meeting on Detection and Attribution of Anthropogenic Climate Change.*
- 32 Hendrix, C. S., 2018: Searching for climate–conflict links. *Nat. Clim. Chang.*, **8**, 190–191,
33 doi:10.1038/s41558-018-0083-3. <http://www.nature.com/articles/s41558-018-0083-3> (Accessed
34 May 22, 2018).
- 35 Herrero, M., and Coauthors, 2013: Biomass use, production, feed efficiencies, and greenhouse gas
36 emissions from global livestock systems. *Proc. Natl. Acad. Sci. U. S. A.*,
37 doi:10.1073/pnas.1308149110.
38 <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=3876224&tool=pmcentrez&renderty>
39 [pe=abstract](http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=3876224&tool=pmcentrez&renderty) (Accessed January 21, 2014).
- 40 HLPE, 2012: *Food security and climate change. A report by the High Level Panel of Experts on Food*
41 *Security and Nutrition of the Committee on World Food Security.* Rome, 102 pp.
42 http://www.fao.org/fileadmin/user_upload/hlpe/hlpe_documents/HLPE_Reports/HLPE-Report-
43 [3-Food_security_and_climate_change-June_2012.pdf](http://www.fao.org/fileadmin/user_upload/hlpe/hlpe_documents/HLPE_Reports/HLPE-Report-) (Accessed March 19, 2018).
- 44 Hochman, Z., D. L. Gobbett, and H. Horan, 2017: Climate trends account for stalled wheat yields in
45 Australia since 1990. *Glob. Chang. Biol.*, **23**, 2071–2081, doi:10.1111/gcb.13604.
46 <http://doi.wiley.com/10.1111/gcb.13604> (Accessed May 17, 2018).
- 47 Huang, J., and F. Ji, 2015: Effects of climate change on phenological trends and seed cotton yields in
48 oasis of arid regions. *Int. J. Biometeorol.*, **59**, 877–888, doi:10.1007/s00484-014-0904-7.

- 1 <http://link.springer.com/10.1007/s00484-014-0904-7> (Accessed April 12, 2019).
- 2 Huitema, D., W. N. Adger, F. Berkhout, E. Massey, D. Mazmanian, S. Munaretto, R. Plummer, and
3 C. C. J. A. M. Termeer, 2016: The governance of adaptation: Choices, reasons, and effects.
4 Introduction to the special feature. *Ecol. Soc.*, **21**, doi:10.5751/ES-08797-210337.
- 5 Hussain, A., G. Rasul, B. Mahapatra, and S. Tuladhar, 2016: Household food security in the face of
6 climate change in the Hindu-Kush Himalayan region. *Food Secur.*, **8**, 921–937,
7 doi:10.1007/s12571-016-0607-5. <http://dx.doi.org/10.1007/s12571-016-0607-5>.
- 8 Ifeanyi-obi, C. C., A. O. Togun, and R. Lamboll, 2016: Influence of Climate Change on Cocoyam
9 Production in Aba Agricultural Zone of Abia State, Nigeria. Springer, Cham, 261–273
10 http://link.springer.com/10.1007/978-3-319-25814-0_18 (Accessed November 4, 2018).
- 11 Iizumi, T., M. Yokozawa, and M. Nishimori, 2011: Probabilistic evaluation of climate change impacts
12 on paddy rice productivity in Japan. *Clim. Change*, **107**, 391–415, doi:10.1007/s10584-010-
13 9990-7. <http://link.springer.com/10.1007/s10584-010-9990-7> (Accessed April 16, 2019).
- 14 IPCC, 2018: *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of*
15 *1.5 °C above pre-industrial levels and related global greenhouse gas emission pathways, in the*
16 *context of strengthening the global response to the threat of climate change.*
17 <http://www.ipcc.ch/report/sr15/>.
- 18 Jat, M. L., and Coauthors, 2016: Climate change and agriculture: Adaptation strategies and mitigation
19 opportunities for food security in South Asia and Latin America. *Advances in Agronomy*, Vol.
20 137 of, 127–235.
- 21 Jayanegara, A., K. A. Sarwono, M. Kondo, H. Matsui, M. Ridla, E. B. Laconi, and Nahrowi, 2018:
22 Use of 3-nitrooxypropanol as feed additive for mitigating enteric methane emissions from
23 ruminants: a meta-analysis. *Ital. J. Anim. Sci.*, **17**, 650–656,
24 doi:10.1080/1828051X.2017.1404945.
- 25 Kelley, C., S. Mohtadi, M. Cane, R. Seager, and Y. Kushnir, 2017: Commentary on the Syria case:
26 Climate as a contributing factor. *Polit. Geogr.*, **60**, 245–247, doi:10.1016/j.polgeo.2017.06.013.
27 <http://dx.doi.org/10.1016/j.polgeo.2017.06.013>.
- 28 Kelley, C. P., S. Mohtadi, M. A. Cane, R. Seager, and Y. Kushnir, 2015: Climate change in the Fertile
29 Crescent and implications of the recent Syrian drought. *Proc. Natl. Acad. Sci.*, **112**, 3241–3246,
30 doi:10.1073/pnas.1421533112. <http://www.pnas.org/lookup/doi/10.1073/pnas.1421533112>.
- 31 Keohane, R. O., and D. G. Victor, 2011: The Regime Complex for Climate Change. *Perspect. Polit.*,
32 **9**, 7–23, doi:10.1017/S1537592710004068.
33 http://www.journals.cambridge.org/abstract_S1537592710004068 (Accessed February 28,
34 2019).
- 35 Khoury, C., O. K., and A. Jarvis, 2014: The Changing Composition of the Global Diet: Implications
36 for CGIAR Research. *Policy Br.*, **18**, 1–6.
- 37 Kyle, G. P., P. Luckow, K. V. Calvin, W. R. Emanuel, M. Nathan, and Y. Zhou, 2011: *GCAM 3.0*
38 *Agriculture and Land Use: Data Sources and Methods*. Richland, WA (United States),.
- 39 Lambek, N., 2019: The UN Committee on World Food Security’s break from the agricultural
40 productivity trap. *Transnatl. Leg. Theory*, 1–15, doi:10.1080/20414005.2018.1573406.
41 <https://www.tandfonline.com/doi/full/10.1080/20414005.2018.1573406> (Accessed March 2,
42 2019).
- 43 Lang, T., and D. Barling, 2012: Food security and food sustainability: reformulating the debate.
44 *Geogr. J.*, **178**, 313–326, doi:10.1111/j.1475-4959.2012.00480.x.
45 <http://doi.wiley.com/10.1111/j.1475-4959.2012.00480.x> (Accessed February 28, 2019).
- 46 Li, T. A. O., and Coauthors, 2015: Uncertainties in predicting rice yield by current crop models under
47 a wide range of climatic conditions. *Glob. Chang. Biol.*, **21**, 1328–1341, doi:10.1111/gcb.12758.

- 1 Li, Z. peng, Y. qiao LONG, P. qin TANG, J. yang TAN, Z. guo LI, W. bin WU, Y. nan HU, and P.
2 YANG, 2017: Spatio-temporal changes in rice area at the northern limits of the rice cropping
3 system in China from 1984 to 2013. *J. Integr. Agric.*, **16**, 360–367, doi:10.1016/S2095-
4 3119(16)61365-5.
- 5 Lombardozzi, D. L., G. B. Bonan, S. Levis, and D. M. Lawrence, 2018: Changes in wood biomass
6 and crop yields in response to projected CO₂, O₃, nitrogen deposition, and climate. *J. Geophys.*
7 *Res. Biogeosciences*.
- 8 López-i-Gelats, F., E. D. G. Fraser, J. F. Morton, and M. G. Rivera-Ferre, 2016: What drives the
9 vulnerability of pastoralists to global environmental change? A qualitative meta-analysis. *Glob.*
10 *Environ. Chang.*, **39**, 258–274, doi:10.1016/J.GLOENVCHA.2016.05.011.
11 <https://www.sciencedirect.com/science/article/abs/pii/S095937801630070X>
12 (Accessed November 3, 2018).
- 13 Lotze-Campen, H., C. Müller, A. Bondeau, S. Rost, A. Popp, and W. Lucht, 2008: Global food
14 demand, productivity growth, and the scarcity of land and water resources: a spatially explicit
15 mathematical programming approach. *Agric. Econ.*, **39**, 325–338, doi:10.1111/j.1574-
16 0862.2008.00336.x.
- 17 Lynch, J., 2019: Availability of disaggregated greenhouse gas emissions from beef cattle production:
18 A systematic review. *Environ. Impact Assess. Rev.*, **76**, 69–78, doi:10.1016/J.EIAR.2019.02.003.
- 19 —, and R. Pierrehumbert, 2019: Climate Impacts of Cultured Meat and Beef Cattle. *Front. Sustain.*
20 *Food Syst.*, **3**, 5, doi:10.3389/fsufs.2019.00005.
- 21 Margulis, M., and J. A. B. Duncan, 2016: Global Food Security Governance : Key Actors, Issues, and
22 Dynamics. *Critical perspectives in food studies*, A. Winson, J. Sumner, and M. Koç, Eds.,
23 Oxford University Press, p. 400 <https://library.wur.nl/WebQuery/wurpubs/510301> (Accessed
24 March 2, 2019).
- 25 Margulis, M. E., 2013: The Regime Complex for Food Security: Implications for the Global Hunger
26 Challenge. *Glob. Gov.*, **19**.
27 [https://heinonline.org/HOL/Page?handle=hein.journals/glogo19&id=55&div=8&collection=jour](https://heinonline.org/HOL/Page?handle=hein.journals/glogo19&id=55&div=8&collection=journals)
28 [nals](https://heinonline.org/HOL/Page?handle=hein.journals/glogo19&id=55&div=8&collection=journals) (Accessed February 27, 2019).
- 29 McKune, S. L., E. C. Borresen, A. G. Young, T. D. Auria Ryley, S. L. Russo, A. Diao Camara, M.
30 Coleman, and E. P. Ryan, 2015: Climate change through a gendered lens: Examining livestock
31 holder food security. *Glob. Food Sec.*, **6**, 1–8, doi:10.1016/J.GFS.2015.05.001.
32 <https://www.sciencedirect.com/science/article/pii/S221191241500022X> (Accessed October 20,
33 2018).
- 34 Medek, D. E., J. Schwartz, and S. S. Myers, 2017: Estimated Effects of Future Atmospheric CO₂
35 Concentrations on Protein Intake and the Risk of Protein Deficiency by Country and Region.
36 *Environ. Health Perspect.*, **125**, doi:10.1289/ehp41.
- 37 Medina, A., A. Akbar, A. Baazeem, A. Rodriguez, and N. Magan, 2017: Climate change, food
38 security and mycotoxins: Do we know enough? *Fungal Biol. Rev.*, **31**, 143–154,
39 doi:10.1016/J.FBR.2017.04.002.
40 <https://www.sciencedirect.com/science/article/pii/S1749461317300179> (Accessed October 1,
41 2018).
- 42 Meng, Q., P. Hou, D. B. Lobell, H. Wang, Z. Cui, F. Zhang, and X. Chen, 2014: The benefits of
43 recent warming for maize production in high latitude China. *Clim. Change*, **122**, 341–349,
44 doi:10.1007/s10584-013-1009-8. <http://link.springer.com/10.1007/s10584-013-1009-8>
45 (Accessed November 4, 2018).
- 46 Meyfroidt, P., 2017: Trade - offs between environment and livelihoods : Bridging the global land use
47 and food security discussions. doi:10.1016/j.gfs.2017.08.001.
- 48 Misselhorn, A., P. Aggarwal, P. Ericksen, P. Gregory, L. Horn-Phathanothai, J. Ingram, and K.
49 Wiebe, 2012: A vision for attaining food security. *Curr. Opin. Environ. Sustain.*, **4**, 7–17,

- 1 doi:10.1016/J.COSUST.2012.01.008.
2 <https://www.sciencedirect.com/science/article/pii/S1877343512000097> (Accessed February 28,
3 2019).
- 4 Mistry, M. N., I. Sue Wing, and E. De Cian, 2017: Simulated vs. empirical weather responsiveness of
5 crop yields: US evidence and implications for the agricultural impacts of climate change.
6 *Environ. Res. Lett.*, **12**, 075007, doi:10.1088/1748-9326/aa788c. <http://stacks.iop.org/1748-9326/12/i=7/a=075007?key=crossref.5e0cd56ab95e1018d6501858e49d9965> (Accessed April
7 16, 2019).
8
- 9 Monkelbaan, J., 2019: Overview of Governance Theories That Are Relevant for the SDGs. Springer,
10 Singapore, 21–48 http://link.springer.com/10.1007/978-981-13-0475-0_2 (Accessed March 2,
11 2019).
- 12 Montanarella, L., and R. Vargas, 2012: Global governance of soil resources as a necessary condition
13 for sustainable development. *Curr. Opin. Environ. Sustain.*, **4**, 559–564,
14 doi:10.1016/J.COSUST.2012.06.007.
15 <https://www.sciencedirect.com/science/article/pii/S1877343512000735> (Accessed April 17,
16 2019).
- 17 Moore, F. C., and D. B. Lobell, 2015: The fingerprint of climate trends on European crop yields.
18 *Proc. Natl. Acad. Sci.*, 201409606.
- 19 Morris, G. P., S. Reis, S. A. Beck, L. E. Fleming, W. N. Adger, T. G. Benton, and M. H. Depledge,
20 2017: Scoping the proximal and distal dimensions of climate change on health and wellbeing.
21 *Environ. Heal.*, **16**, 116, doi:10.1186/s12940-017-0329-y.
22 <https://ehjournal.biomedcentral.com/articles/10.1186/s12940-017-0329-y> (Accessed January 8,
23 2018).
- 24 Müller, C., J. Elliott, and A. Levermann, 2014: Food security: Fertilizing hidden hunger. *Nat. Clim.*
25 *Chang.*, **4**, 540–541, doi:10.1038/nclimate2290.
26 <http://www.nature.com/doi/10.1038/nclimate2290> (Accessed December 6, 2017).
- 27 Myers, S. S., and Coauthors, 2014: Increasing CO2 threatens human nutrition. *Nature*, **510**, 139–142,
28 doi:10.1038/nature13179. <http://www.nature.com/doi/10.1038/nature13179>.
- 29 —, K. R. Wessells, I. Kloog, A. Zanobetti, and J. Schwartz, 2015: Effect of increased
30 concentrations of atmospheric carbon dioxide on the global threat of zinc deficiency: A
31 modelling study. *Lancet Glob. Heal.*, **3**, e639–e645, doi:10.1016/S2214-109X(15)00093-5.
- 32 —, M. R. Smith, S. Guth, C. D. Golden, B. Vaitla, N. D. Mueller, A. D. Dangour, and P. Huybers,
33 2017: Climate Change and Global Food Systems: Potential Impacts on Food Security and
34 Undernutrition. *Annu. Rev. Public Health*, **38**, 259–277, doi:10.1146/annurev-publhealth-
35 031816-044356. [http://www.annualreviews.org/doi/10.1146/annurev-publhealth-031816-
36 044356](http://www.annualreviews.org/doi/10.1146/annurev-publhealth-031816-044356).
- 37 Newbold, T., and Coauthors, 2015: Global effects of land use on local terrestrial biodiversity. *Nature*,
38 **520**, 45–50,
39 doi:10.1038/nature14324 [http://www.nature.com/nature/journal/v520/n7545/abs/nature14324.ht
40 ml#supplementary-information](http://www.nature.com/nature/journal/v520/n7545/abs/nature14324.html#supplementary-information). <http://dx.doi.org/10.1038/nature14324>.
- 41 Newell, P., O. Taylor, and C. Touni, 2018: Governing Food and Agriculture in a Warming World.
42 *Glob. Environ. Polit.*, **18**, 53–71, doi:10.1162/glep_a_00456.
43 https://www.mitpressjournals.org/doi/abs/10.1162/glep_a_00456 (Accessed February 27, 2019).
- 44 Obersteiner, M., and Coauthors, 2016: Assessing the land resource-food price nexus of the
45 Sustainable Development Goals. *Sci. Adv.*, **2**, e1501499–e1501499,
46 doi:10.1126/sciadv.1501499. <http://advances.sciencemag.org/cgi/doi/10.1126/sciadv.1501499>.
- 47 de Oliveira Silva, R., L. G. Barioni, J. A. J. Hall, M. Folegatti Matsuura, T. Zanett Albertini, F. A.
48 Fernandes, and D. Moran, 2016: Increasing beef production could lower greenhouse gas
49 emissions in Brazil if decoupled from deforestation. *Nat. Clim. Chang.*, **6**, 493–497,

- 1 doi:10.1038/nclimate2916. <http://www.nature.com/articles/nclimate2916>.
- 2 Ongoro, E., and W. Ogara, 2012: Impact of climate change and gender roles in community
3 adaptation: A case study of pastoralists in Samburu East District, Kenya. *Int. J. Biodivers.*
4 *Conserv.*, **42**, 78–89.
5 [https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_r](https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_roles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Kenya)
6 [oles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Keny](https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_roles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Kenya)
7 [a](https://www.researchgate.net/publication/267785754_Impact_of_climate_change_and_gender_roles_in_community_adaptation_A_case_study_of_pastoralists_in_Samburu_East_District_Kenya) (Accessed March 24, 2019).
- 8 Onyeneke, R. U., 2018: *International journal of biosciences, agriculture and technology : IJBSAT*. A.
9 Parbhu Britto, 1-7 pp. <https://www.cabdirect.org/cabdirect/abstract/20183191054> (Accessed
10 November 4, 2018).
- 11 Orr, B. J., and Coauthors, 2017: *Scientific Conceptual Framework for Land Degradation Neutrality. A*
12 *Report of the Science-Policy Interface*. Bonn, Germany, 98 pp.
- 13 Ostrom, E., 2010: Polycentric systems for coping with collective action and global environmental
14 change. *Glob. Environ. Chang.*, **20**, 550–557, doi:10.1016/J.GLOENVCHA.2010.07.004.
15 <https://www.sciencedirect.com/science/article/pii/S0959378010000634> (Accessed May 31,
16 2018).
- 17 Palm, C. A., and Coauthors, 2010: Identifying potential synergies and trade-offs for meeting food
18 security and climate change objectives in sub-Saharan Africa. *Proc. Natl Acad. Sci.*, **107**, 19661,
19 doi:10.1073/pnas.0912248107.
- 20 Paterson, R. R. M., and N. Lima, 2011: Further mycotoxin effects from climate change. *Food Res.*
21 *Int.*, **44**, 2555–2566, doi:10.1016/J.FOODRES.2011.05.038.
22 <https://www.sciencedirect.com/science/article/pii/S0963996911003541> (Accessed April 23,
23 2018).
- 24 Paulot, F., and D. J. Jacob, 2014: Hidden cost of U.S. agricultural exports: particulate matter from
25 ammonia emissions. *Env. Sci Technol*, **48**, 903–908, doi:10.1021/es4034793.
- 26 Pereira, L. M., and S. Ruysenaar, 2012: Moving from traditional government to new adaptive
27 governance: the changing face of food security responses in South Africa. *Food Secur.*, **4**, 41–
28 58, doi:10.1007/s12571-012-0164-5. <http://link.springer.com/10.1007/s12571-012-0164-5>
29 (Accessed February 28, 2019).
- 30 Persson, U. M., S. Henders, and C. Cederberg, 2014: A method for calculating a land-use change
31 carbon footprint (LUC-CFP) for agricultural commodities - applications to Brazilian beef and
32 soy, Indonesian palm oil. *Glob. Chang. Biol.*, **20**, 3482–3491, doi:10.1111/gcb.12635.
- 33 Pikaar, I., S. Matassa, K. Rabaey, B. Laycock, N. Boon, and W. Verstraete, 2018: The Urgent Need to
34 Re-engineer Nitrogen-Efficient Food Production for the Planet. *Managing Water, Soil and*
35 *Waste Resources to Achieve Sustainable Development Goals*, Springer International Publishing,
36 Cham, 35–69 http://link.springer.com/10.1007/978-3-319-75163-4_3 (Accessed May 31, 2018).
- 37 Popp, A., and Coauthors, 2014: Land-use protection for climate change mitigation. *Nat. Clim. Chang.*,
38 **4**, 1095–1098, doi:10.1038/nclimate2444.
- 39 Porter, J. R., L. Xie, A. J. Challinor, K. Cochrane, S. M. Howden, M. M. Iqbal, D. B. Lobell, and M.
40 I. Travasso, 2014: Food security and food production systems. *Clim. Chang. 2014 Impacts,*
41 *Adapt. Vulnerability. Part A Glob. Sect. Asp. Contrib. Work. Gr. II to Fifth Assess. Rep.*
42 *Intergov. Panel Clim. Chang.*, **2**, 485–533, doi:10.1111/j.1728-4457.2009.00312.x.
- 43 Potopová, V., P. Zahradníček, P. Štěpánek, L. Türkott, A. Farda, and J. Soukup, 2017: The impacts of
44 key adverse weather events on the field-grown vegetable yield variability in the Czech Republic
45 from 1961 to 2014. *Int. J. Climatol.*, **37**, 1648–1664, doi:10.1002/joc.4807.
46 <http://doi.wiley.com/10.1002/joc.4807> (Accessed November 4, 2018).
- 47 Pradhan, P., L. Costa, D. Rybski, W. Lucht, and J. P. Kropp, 2017: A Systematic Study of Sustainable
48 Development Goal (SDG) Interactions. *Earth's Futur.*, **5**, 1169–1179,

- 1 doi:10.1002/2017EF000632.
- 2 Ramankutty, N., Z. Mehrabi, K. Waha, L. Jarvis, C. Kremen, M. Herrero, and L. H. Rieseberg, 2018:
3 Trends in Global Agricultural Land Use: Implications for Environmental Health and Food
4 Security. *Annu. Rev. Plant Biol.*, **69**, 789–815, doi:10.1146/annurev-arplant-042817-040256.
5 <http://www.annualreviews.org/doi/10.1146/annurev-arplant-042817-040256> (Accessed March 4,
6 2019).
- 7 Rasul, G., A. Saboor, P. C. Tiwari, A. Hussain, N. Ghosh, and G. B. Chettri, 2019: Food and
8 Nutritional Security in the Hindu Kush Himalaya: Unique Challenges and Niche Opportunities.
9 *The Hindu Kush Himalaya Assessment: mountains, climate change, sustainability and people.*,
10 P. Wester, A. Mishra, A. Mukherji, and A.B. Shrestha, Eds., SPRINGER, Dordrecht
11 <https://www.springer.com/us/book/9783319922874> (Accessed October 29, 2018).
- 12 Reisinger, A., and H. Clark, 2018: How much do direct livestock emissions actually contribute to
13 global warming? *Glob. Chang. Biol.*, **24**, 1749–1761, doi:10.1111/gcb.13975.
14 <http://doi.wiley.com/10.1111/gcb.13975> (Accessed March 19, 2018).
- 15 Richardson, J., A. Nichols, and T. Henry, 2012: Do transition towns have the potential to promote
16 health and well-being? A health impact assessment of a transition town initiative. *Public Health*,
17 **126**, 982–989, doi:10.1016/J.PUHE.2012.07.009.
18 <https://www.sciencedirect.com/science/article/pii/S0033350612002739> (Accessed February 26,
19 2019).
- 20 Rivera-Ferre, M., M. Ortega-Cerdà, J. Baumgärtner, M. G. Rivera-Ferre, M. Ortega-Cerdà, and J.
21 Baumgärtner, 2013: Rethinking Study and Management of Agricultural Systems for Policy
22 Design. *Sustainability*, **5**, 3858–3875, doi:10.3390/su5093858. [http://www.mdpi.com/2071-](http://www.mdpi.com/2071-1050/5/9/3858)
23 [1050/5/9/3858](http://www.mdpi.com/2071-1050/5/9/3858) (Accessed October 31, 2018).
- 24 Rivera-Ferre, M. G., 2012: Framing of Agri-food Research Affects the Analysis of Food Security:
25 The Critical Role of the Social Sciences. *Int. J. Sociol. Agric. Food*, **19**, 162–175.
26 <http://ijsaf.org/contents/19-2/rivera-ferre/index.html> (Accessed December 8, 2017).
- 27 Robinson, S., and Coauthors, 2015: The international model for policy analysis of agricultural
28 commodities and trade (IMPACT): Model description for version 3.
- 29 Rosenzweig, C., and D. Hillel, 2015: *Handbook of Climate Change and Agroecosystems: The*
30 *Agricultural Model Intercomparison and Improvement Project (AgMIP) Integrated Crop and*
31 *Economic Assessments*. First. World Scientific,.
- 32 —, and Coauthors, 2014a: Assessing agricultural risks of climate change in the 21st century in a
33 global gridded crop model intercomparison. *Proc. Natl. Acad. Sci.*, **111**, 3268–3273,
34 doi:10.1073/pnas.1222463110. <http://www.pnas.org/lookup/doi/10.1073/pnas.1222463110>.
- 35 —, and Coauthors, 2014b: Assessing agricultural risks of climate change in the 21st century in a
36 global gridded crop model intercomparison. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 3268–3273,
37 doi:10.1073/pnas.1222463110.
38 <http://www.pubmedcentral.nih.gov/articlerender.fcgi?artid=3948251&tool=pmcentrez&rendertype=abstract>.
- 39
- 40 —, and Coauthors, 2017: Coordinating AgMIP data and models across global and regional scales
41 for 1.5 °C and 2.0 °C assessments. *Phil. Trans. R. Soc.*, doi:10.1098/rsta.2016.0455.
- 42 —, W. D. Solecki, P. Romero-Lankao, S. Mehrotra, S. Dhakal, and S. A. Ibrahim, 2018: *Climate*
43 *Change and Cities: Second Assessment Report of the Urban Climate Change Research Network*.
44 Cambridge University Press,.
- 45 Ruane, A. C., and Coauthors, 2018: Biophysical and economic implications for agriculture of +1.5°
46 and +2.0°C global warming using AgMIP Coordinated Global and Regional Assessments. *Clim.*
47 *Res.*, **76**, 17–39, doi:https://doi.org/10.3354/cr01520. [https://www.int-](https://www.int-res.com/abstracts/cr/v76/n1/p17-39/)
48 [res.com/abstracts/cr/v76/n1/p17-39/](https://www.int-res.com/abstracts/cr/v76/n1/p17-39/).

- 1 Sage, C., 2014: The transition movement and food sovereignty: From local resilience to global
2 engagement in food system transformation. *J. Consum. Cult.*, **14**, 254–275,
3 doi:10.1177/1469540514526281. <http://journals.sagepub.com/doi/10.1177/1469540514526281>
4 (Accessed February 26, 2019).
- 5 Sapkota, T. B., M. L. Jat, J. P. Aryal, R. K. Jat, and A. Khatri-Chhetri, 2015: Climate change
6 adaptation, greenhouse gas mitigation and economic profitability of conservation agriculture:
7 Some examples from cereal systems of Indo-Gangetic Plains. *J. Integr. Agric.*, **14**, 1524–1533,
8 doi:10.1016/S2095-3119(15)61093-0.
- 9 Sapkota, T. B., J. P. Aryal, A. Khatri-chhetri, P. B. Shirsath, P. Arumugam, and C. M. Stirling, 2017:
10 Identifying high-yield low-emission pathways for the cereal production in South Asia. *Mitig.*
11 *Adapt. Strateg. Glob. Chang.*, doi:10.1007/s11027-017-9752-1.
- 12 Schiff, R., 2008: The Role of Food Policy Councils in Developing Sustainable Food Systems. *J.*
13 *Hunger Environ. Nutr.*, **3**, 206–228, doi:10.1080/19320240802244017.
14 <http://www.tandfonline.com/doi/abs/10.1080/19320240802244017> (Accessed February 26,
15 2019).
- 16 Schmidhuber, J., and F. N. Tubiello, 2007: Global food security under climate change. *Proc. Natl.*
17 *Acad. Sci. U. S. A.*, **104**, 19703–19708, doi:10.1073/pnas.0701976104.
18 <http://www.pnas.org/content/104/50/19703.short>.
- 19 Schmidinger, K., and E. Stehfest, 2012: Including CO2 implications of land occupation in LCAs—
20 method and example for livestock products. *Int. J. Life Cycle Assess.*, **17**, 962–972,
21 doi:10.1007/s11367-012-0434-7.
- 22 Schmidt, J. H., B. P. Weidema, and M. Brandão, 2015: A framework for modelling indirect land use
23 changes in Life Cycle Assessment. *J. Clean. Prod.*, **99**, 230–238,
24 doi:10.1016/J.JCLEPRO.2015.03.013.
- 25 Schweiger, O., and Coauthors, 2010: Multiple stressors on biotic interactions: how climate change
26 and alien species interact to affect pollination. *Biol. Rev.*, **85**, 777–795, doi:10.1111/j.1469-
27 185X.2010.00125.x.
- 28 Selby, J., O. S. Dahi, C. Fröhlich, and M. Hulme, 2017: Climate change and the Syrian civil war
29 revisited. *Polit. Geogr.*, **60**, 232–244, doi:10.1016/J.POLGEO.2017.05.007.
30 <https://www.sciencedirect.com/science/article/pii/S0962629816301822> (Accessed May 22,
31 2018).
- 32 Smetana, S., A. Mathys, A. Knoch, and V. Heinz, 2015: Meat alternatives: life cycle assessment of
33 most known meat substitutes. *Int. J. Life Cycle Assess.*, **20**, 1254–1267, doi:10.1007/s11367-
34 015-0931-6.
- 35 Smith, M. R., and S. S. Myers, 2018: Impact of anthropogenic CO2 emissions on global human
36 nutrition. *Nat. Clim. Chang.*, **8**, 834–839, doi:10.1038/s41558-018-0253-3.
37 <http://dx.doi.org/10.1038/s41558-018-0253-3>.
- 38 Smith, M. R., C. D. Golden, and S. S. Myers, 2017: Potential rise in iron deficiency due to future
39 anthropogenic carbon dioxide emissions. *GeoHealth*, 1–10, doi:10.1002/2016GH000018.
40 <http://doi.wiley.com/10.1002/2016GH000018>.
- 41 Springmann, M., and Coauthors, 2018a: Options for keeping the food system within environmental
42 limits. *Nature*, doi:10.1038/s41586-018-0594-0. <https://doi.org/10.1038/s41586-018-0594-0>.
- 43 —, K. Wiebe, D. Mason-D’Croz, T. B. Sulser, M. Rayner, and P. Scarborough, 2018b: Health and
44 nutritional aspects of sustainable diet strategies and their association with environmental
45 impacts: a global modelling analysis with country-level detail. *Lancet. Planet. Heal.*, **2**, e451–
46 e461, doi:10.1016/S2542-5196(18)30206-7. <http://www.ncbi.nlm.nih.gov/pubmed/30318102>
47 (Accessed November 5, 2018).
- 48 Stehfest, E., D. van Vuuren, L. Bouwman, and T. Kram, 2014: *Integrated assessment of global*

- 1 *environmental change with IMAGE 3.0: Model description and policy applications*. Hague,.
- 2 Tamea, S., F. Laio, and L. Ridolfi, 2016: Global effects of local food-production crises: a virtual
3 water perspective. *Sci. Rep.*, **6**, doi:10.1038/srep18803.
- 4 Tao, F., and Coauthors, 2014: Responses of wheat growth and yield to climate change in different
5 climate zones of China, 1981–2009. *Agric. For. Meteorol.*, **189–190**, 91–104,
6 doi:10.1016/J.AGRFORMET.2014.01.013.
7 <https://www.sciencedirect.com/science/article/pii/S0168192314000227?via%3Dihub> (Accessed
8 April 12, 2019).
- 9 Tariq, M., and Coauthors, 2018: The impact of climate warming and crop management on phenology
10 of sunflower-based cropping systems in Punjab, Pakistan. *Agric. For. Meteorol.*, **256–257**, 270–
11 282, doi:10.1016/J.AGRFORMET.2018.03.015.
12 <https://www.sciencedirect.com/science/article/pii/S0168192318300984> (Accessed November 4,
13 2018).
- 14 Taylor Aiken, G., 2015: (Local-) community for global challenges: carbon conversations, transition
15 towns and governmental elisions. *Local Environ.*, **20**, 764–781,
16 doi:10.1080/13549839.2013.870142.
17 <http://www.tandfonline.com/doi/full/10.1080/13549839.2013.870142> (Accessed February 26,
18 2019).
- 19 Teh, L. C. L., and U. R. Sumaila, 2013: Contribution of marine fisheries to worldwide employment.
20 *Fish Fish.*, **14**, 77–88, doi:10.1111/j.1467-2979.2011.00450.x.
21 <http://doi.wiley.com/10.1111/j.1467-2979.2011.00450.x> (Accessed March 26, 2018).
- 22 Termeer, C. J. A. M., S. Drimie, J. Ingram, L. Pereira, and M. J. Whittingham, 2018: A diagnostic
23 framework for food system governance arrangements: The case of South Africa. *NJAS -*
24 *Wageningen J. Life Sci.*, **84**, 85–93, doi:10.1016/J.NJAS.2017.08.001.
25 <https://www.sciencedirect.com/science/article/pii/S157352141730012X> (Accessed May 4,
26 2018).
- 27 Thompson, B., M. J. Cohen, and J. Meerman, 2012: *The Impact of Climate Change and Bioenergy on*
28 *Nutrition*. Springer Netherlands, Dordrecht, 103 pp.
29 http://www.springerlink.com/index/10.1007/978-94-007-0110-6_1 (Accessed December 8,
30 2017).
- 31 Thornton, P. K., and M. Herrero, 2010: Potential for reduced methane and carbon dioxide emissions
32 from livestock and pasture management in the tropics. *Proc. Natl. Acad. Sci. U. S. A.*, **107**,
33 19667–19672, doi:10.1073/pnas.0912890107.
- 34 Thow, A. M., and Coauthors, 2018: How can health, agriculture and economic policy actors work
35 together to enhance the external food environment for fruit and vegetables? A qualitative policy
36 analysis in India. *Food Policy*, **77**, 143–151, doi:10.1016/J.FOODPOL.2018.04.012.
37 <https://www.sciencedirect.com/science/article/pii/S0306919217304372> (Accessed March 4,
38 2019).
- 39 Tirado, M. C., and J. Meerman, 2012: Climate change and food and nutrition security. *The Impact of*
40 *climate change and bioenergy on nutrition*, Springer, 43–60.
- 41 Tittensor, D. P. et al., 2017: A protocol for the intercomparison of marine fishery and ecosystem
42 models: Fish-MIP v1.0. *Geosci. Model Dev.*, doi:[https://www.geosci-model-dev-](https://www.geosci-model-dev-discuss.net/gmd-2017-209/)
43 [discuss.net/gmd-2017-209/](https://www.geosci-model-dev-discuss.net/gmd-2017-209/).
- 44 Tittonell, P., M. T. van Wijk, M. Herrero, M. C. Rufino, N. de Ridder, and K. E. Giller, 2009: Beyond
45 resource constraints - Exploring the biophysical feasibility of options for the intensification of
46 smallholder crop-livestock systems in Vihiga district, Kenya. *Agric. Syst.*, **101**,
47 doi:10.1016/j.agsy.2009.02.003.
- 48 Toreti, A., and Coauthors, Elevated CO₂ and agriculture: uncertainties, opportunities and policy
49 implications. *Nat. Clim. Chang.*, **Forthcomin.**

- 1 Tukker, A., R. A. Goldbohm, A. de Koning, M. Verheijden, R. Kleijn, O. Wolf, I. Pérez-Domínguez,
2 and J. M. Rueda-Cantucho, 2011: Environmental impacts of changes to healthier diets in
3 Europe. *Ecol. Econ.*, **70**, 1776–1788, doi:10.1016/J.ECOLECON.2011.05.001.
- 4 Uddin, S., M. Löw, S. Parvin, G. J. Fitzgerald, S. Tausz-Posch, R. Armstrong, and M. Tausz, 2018:
5 Yield of canola (*Brassica napus* L.) benefits more from elevated CO₂ when access to deeper soil
6 water is improved. *Environ. Exp. Bot.*, **155**, 518–528, doi:10.1016/J.ENVEXPBOT.2018.07.017.
7 <https://www.sciencedirect.com/science/article/pii/S0098847218306324> (Accessed February 27,
8 2019).
- 9 UNCCD, 2017: *Global Land Outlook*. Bonn, Germany,.
- 10 Velthof, G. L., and J. Mosquera, 2011: Calculation of nitrous oxide emission from agriculture in the
11 Netherlands.
- 12 Vermeulen, S. J., B. M. Campbell, and J. S. I. Ingram, 2012: Climate change and food systems. *Annu.*
13 *Rev. Environ. Resour.*, **37**, 195–222, doi:10.1146/annurev-environ-020411-130608.
- 14 Wang, E., and Coauthors, 2017: The uncertainty of crop yield projections is reduced by improved
15 temperature response functions. *Nat. Plants*, **3**, 17102, doi:10.1038/nplants.2017.102.
16 <http://www.nature.com/articles/nplants2017102>.
- 17 Wellesley, L., F. Preston, J. Lehne, and R. Bailey, 2017: Chokepoints in global food trade: Assessing
18 the risk. *Res. Transp. Bus. Manag.*, **25**, 15–28, doi:10.1016/J.RTBM.2017.07.007.
19 <https://www.sciencedirect.com/science/article/pii/S2210539517300172> (Accessed January 8,
20 2018).
- 21 West, P. C., and Coauthors, 2014: Leverage points for improving global food security and the
22 environment. *Science* (80-.), **345**, 325–328, doi:10.1126/science.1246067.
23 <http://www.sciencemag.org/cgi/doi/10.1126/science.1246067>.
- 24 Wiebe, K., and Coauthors, 2015: Climate change impacts on agriculture in 2050 under a range of
25 plausible socioeconomic and emissions scenarios. *Environ. Res. Lett.*, **10**, 085010,
26 doi:10.1088/1748-9326/10/8/085010. [http://stacks.iop.org/1748-
27 9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b](http://stacks.iop.org/1748-9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b).
- 28 Willenbockel, D., 2012: *Extreme Weather Events and Crop Price Spikes in a Changing Climate:*
29 *Illustrative global simulation scenarios*. [https://www.oxfam.org/sites/www.oxfam.org/files/rr-
30 extreme-weather-events-crop-price-spikes-05092012-en.pdf](https://www.oxfam.org/sites/www.oxfam.org/files/rr-extreme-weather-events-crop-price-spikes-05092012-en.pdf).
- 31 Willett, W., and Coauthors, 2019: Food in the Anthropocene: the EAT-Lancet Commission on healthy
32 diets from sustainable food systems. *Lancet (London, England)*, **393**, 447–492,
33 doi:10.1016/S0140-6736(18)31788-4. <http://www.ncbi.nlm.nih.gov/pubmed/30660336>
34 (Accessed March 4, 2019).
- 35 Wise, M., and K. Calvin, 2011: *GCAM3.0 Agriculture and Land Use: Technical Description of*
36 *Modeling Approach*.
- 37 Wittman, H., M. J. Chappell, D. J. Abson, R. B. Kerr, J. Blesh, J. Hanspach, I. Perfecto, and J.
38 Fischer, 2017: A social–ecological perspective on harmonizing food security and biodiversity
39 conservation. *Reg. Environ. Chang.*, **17**, 1291–1301, doi:10.1007/s10113-016-1045-9.
- 40 Woltjer, G., A. Kavallari, H. Van Meijl, J. Powell, M. Rutten, and L. Shutes, 2014: *The MAGNET*
41 *model*. 148 pp.
- 42 Yan, M.-J., J. Humphreys, and N. M. Holden, 2011: An evaluation of life cycle assessment of
43 European milk production. *J. Environ. Manage.*, **92**, 372–379,
44 doi:10.1016/J.JENVMAN.2010.10.025.
- 45 Zech, K. M., and U. A. Schneider, 2019: Carbon leakage and limited efficiency of greenhouse gas
46 taxes on food products. *J. Clean. Prod.*, **213**, 99–103, doi:10.1016/J.JCLEPRO.2018.12.139.
- 47 Zhao, C., and Coauthors, 2017: Temperature increase reduces global yields of major crops in four

- 1 independent estimates. *Proc. Natl. Acad. Sci. U. S. A.*, **114**, 9326–9331,
2 doi:10.1073/pnas.1701762114. <http://www.ncbi.nlm.nih.gov/pubmed/28811375> (Accessed
3 October 1, 2018).
- 4 Zhu, C., and Coauthors, 2018: Carbon dioxide (CO₂) levels this century will alter the protein,
5 micronutrients, and vitamin content of rice grains with potential health consequences for the
6 poorest rice-dependent countries. *Sci. Adv.*, **4**.
7 <http://advances.sciencemag.org/content/4/5/eaq1012.abstract>.
- 8 Zimmermann, A., H. Webber, G. Zhao, F. Ewert, J. Kros, J. Wolf, W. Britz, and W. De Vries, 2017:
9 Climate change impacts on crop yields, land use and environment in response to crop sowing
10 dates and thermal time requirements. *Agric. Syst.*, **157**, 81–92, doi:10.1016/j.agsy.2017.07.007.

1 **Chapter 6: Interlinkages between Desertification, Land**
2 **Degradation, Food Security and GHG fluxes:**
3 **synergies, trade-offs and Integrated Response**
4 **Options**

5
6 **Coordinating Lead Authors:** Pete Smith (United Kingdom), Johnson Nkem (Cameroon), Katherine
7 Calvin (The United States of America)

8 **Lead Authors:** Donovan Campbell (Jamaica), Francesco Cherubini (Norway/Italy), Giacomo Grassi
9 (Italy/EU), Vladimir Korotkov (The Russian Federation), Anh Le Hoang (Viet Nam), Shuaib Lwasa
10 (Uganda), Pamela McElwee (The United States of America), Ephraim Nkonya (Tanzania), Nobuko
11 Saigusa (Japan), Jean-Francois Soussana (France), Miguel Angel Taboada (Argentina)

12 **Contributing Authors:** Cristina Arias-Navarro (Spain), Otavio Cavalett (Brazil), Annette Cowie
13 (Australia), Joanna House (United Kingdom), Daniel Huppmann (Austria), Jagdish Krishnaswamy
14 (India), Alexander Popp (Germany), Stephanie Roe (The Philippines/The United States of America),
15 Raphael Slade (United Kingdom), Lindsay Stringer (United Kingdom), Matteo Vizzarri (Italy)

16 **Review Editors:** Amjad Abdulla (Maldives), Ian Noble (Australia), Yoshiki Yamagata (Japan), Taha
17 Zatari (Saudi Arabia)

18 **Chapter Scientists:** Frances Manning (United Kingdom), Dorothy Nampanzira (Uganda)

19 **Date of Draft:** 07/08/2019

20

21

| | | |
|----|---|-------|
| 1 | Table of Contents | |
| 2 | Chapter 6: Interlinkages between Desertification, Land Degradation, Food Security and GHG | |
| 3 | fluxes: synergies, trade-offs and Integrated Response Options | 6-1 |
| 4 | Executive summary | 6-3 |
| 5 | 6.1 Introduction..... | 6-7 |
| 6 | 6.1.1 Context of this chapter | 6-7 |
| 7 | 6.1.2 Framing social challenges and acknowledging enabling factors | 6-7 |
| 8 | 6.1.3 Challenges and response options in current and historical interventions..... | 6-10 |
| 9 | 6.1.4 Challenges represented in future scenarios | 6-18 |
| 10 | 6.2 Response options, co-benefits and adverse side-effects across the land challenges | 6-21 |
| 11 | 6.2.1 Integrated response options based on land management | 6-25 |
| 12 | 6.2.2 Integrated response options based on value chain management | 6-39 |
| 13 | 6.2.3 Integrated response options based on risk management | 6-39 |
| 14 | Cross-Chapter Box 7: Bioenergy and Bioenergy with Carbon Dioxide Capture and Storage | |
| 15 | (BECCS) in mitigation scenarios | 6-48 |
| 16 | 6.3 Potentials for addressing the land challenges..... | 6-51 |
| 17 | 6.3.1 Potential of the integrated response options for delivering mitigation | 6-51 |
| 18 | 6.3.2 Potential of the integrated response options for delivering adaptation | 6-60 |
| 19 | 6.3.3 Potential of the integrated response options for addressing desertification | 6-68 |
| 20 | 6.3.4 Potential of the integrated response options for addressing land degradation | 6-74 |
| 21 | 6.3.5 Potential of the integrated response options for addressing food security | 6-80 |
| 22 | 6.3.6 Summarising the potential of the integrated response options across mitigation, | |
| 23 | adaptation, desertification land degradation and food security..... | 6-89 |
| 24 | 6.4 Managing interactions and interlinkages | 6-102 |
| 25 | 6.4.1 Feasibility of the integrated response options with respect to costs, barriers, saturation | |
| 26 | and reversibility | 6-102 |
| 27 | 6.4.2 Sensitivity of the Integrated Response Options to climate change impacts..... | 6-111 |
| 28 | Cross-Chapter Box 8: Ecosystem services and Nature’s Contributions to People, and their relation | |
| 29 | to the land-climate system..... | 6-113 |
| 30 | 6.4.3 Impacts of integrated response options on Nature’s Contributions to People and the UN | |
| 31 | Sustainable Development Goals | 6-117 |
| 32 | 6.4.4 Opportunities for implementation of Integrated Response Options..... | 6-130 |
| 33 | Cross-Chapter Box 9: Illustrative Climate and Land Pathways..... | 6-142 |
| 34 | 6.4.5 Potential Consequences of Delayed Action | 6-147 |
| 35 | Frequently Asked Questions | 6-150 |
| 36 | Appendix to Chapter 6: Interlinkages between Desertification, Land Degradation, Food Security and | |
| 37 | GHG fluxes: synergies, trade-offs and Integrated Response Options..... | 6-152 |
| 38 | Supplementary Information for Section 6.4.1 | 6-152 |

| | | |
|---|--|-------|
| 1 | Supplementary Information for Section 6.4.3..... | 6-170 |
| 2 | Supplementary Information for Section 6.4.4..... | 6-204 |
| 3 | References:..... | 6-206 |

4

5 **Executive summary**

6 **The land challenges, in the context of this report, are climate change mitigation, adaptation,**
7 **desertification, land degradation, and food security.** The chapter also discusses implications for
8 Nature’s Contributions to People (NCP), including biodiversity and water, and sustainable
9 development, by assessing intersections with the Sustainable Development Goals (SDGs). The
10 chapter assesses response options that could be used to address these challenges. These response
11 options were derived from the previous chapters and fall into three broad categories (land
12 management, value chain, and risk management).

13 **The land challenges faced today vary across regions; climate change will increase challenges in**
14 **the future, while socioeconomic development could either increase or decrease challenges (*high***
15 ***confidence*).** Increases in biophysical impacts from climate change can worsen desertification, land
16 degradation, and food insecurity (*high confidence*). Additional pressures from socioeconomic
17 development could further exacerbate these challenges; however, the effects are scenario dependent.
18 Scenarios with increases in income and reduced pressures on land can lead to reductions in food
19 insecurity; however, all assessed scenarios result in increases in water demand and water scarcity
20 (*medium confidence*). {6.1}

21 **The applicability and efficacy of response options are region and context specific; while many**
22 **value chain and risk management options are potentially broadly applicable, many land**
23 **management options are applicable on less than 50% of the ice-free land surface (*high***
24 ***confidence*).** Response options are limited by land type, bioclimatic region, or local food system
25 context (*high confidence*). Some response options produce adverse side-effects only in certain regions
26 or contexts; for example, response options that use freshwater may have no adverse side effects in
27 regions where water is plentiful, but large adverse side effects in regions where water is scarce (*high*
28 *confidence*). Response options with biophysical climate effects (e.g., afforestation, reforestation) may
29 have different effects on local climate depending on where they are implemented (*medium*
30 *confidence*). Regions with more challenges have fewer response options available for implementation
31 (*medium confidence*). {6.1, 6.2, 6.3, 6.4}

32 **Nine options deliver medium to large benefits for all five land challenges (*high confidence*).** The
33 options with medium to large benefits for all challenges are increased food productivity, improved
34 cropland management, improved grazing land management, improved livestock management,
35 agroforestry, improved forest management, increased soil organic carbon content, fire management
36 and reduced post-harvest losses. A further two options, dietary change and reduced food waste, have
37 no global estimates for adaptation but have medium to large benefits for all other challenges (*high*
38 *confidence*). {6.3, 6.4}

39 **Five options have large mitigation potential (> 3 GtCO₂e yr⁻¹) without adverse impacts on the**
40 **other challenges (*high confidence*).** These are increased food productivity, reduced deforestation and
41 degradation, increased soil organic carbon content, fire management and reduced post-harvest losses.
42 Two further options with large mitigation potential, dietary change and reduced food waste, have no
43 global estimates for adaptation but show no negative impacts across the other challenges. Five
44 options: improved cropland management, improved grazing land managements, agroforestry,

1 integrated water management and forest management, have moderate mitigation potential, with no
2 adverse impacts on the other challenges (*high confidence*). {6.3.6}

3 **Sixteen response options have large adaptation potential (>25 million people benefit), without**
4 **adverse side-effects on other land challenges (*high confidence*).** These are increased food
5 productivity, improved cropland management, agroforestry, agricultural diversification, improved
6 forest management, increased soil organic carbon content, reduced landslides and natural hazards,
7 restoration and reduced conversion of coastal wetlands, reduced post-harvest losses, sustainable
8 sourcing, management of supply chains, improved food processing and retailing, improved energy use
9 in food systems, livelihood diversification, use of local seeds, and disaster risk management (*high*
10 *confidence*). Some options (such as enhanced urban food systems or management of urban sprawl)
11 may not provide large global benefits but may have significant positive local effects without adverse
12 effects (*high confidence*). {6.3, 6.4}

13 **Seventeen of 40 options deliver co-benefits or no adverse side-effects for the full range of NCPs**
14 **and SDGs; only three options (afforestation, bioenergy and BECCS and some types of risk**
15 **sharing instruments, such as insurance) have potentially adverse side-effects for five or more**
16 **NCPs or SDGs (*medium confidence*).** The 17 options with co-benefits and no adverse side-effects
17 include most agriculture- and soil-based land management options, many ecosystem-based land
18 management options, improved forest management, reduced post-harvest losses, sustainable sourcing,
19 improved energy use in food systems, and livelihood diversification (*medium confidence*). Some of
20 the synergies between response options and SDGs include positive poverty reduction impacts from
21 activities like improved water management or improved management of supply chains. Examples of
22 synergies between response options and NCPs include positive impacts on habitat maintenance from
23 activities like invasive species management and agricultural diversification. However, many of these
24 synergies are not automatic, and are dependent on well-implemented activities requiring institutional
25 and enabling conditions for success. {6.4}

26 **Most response options can be applied without competing for available land; however, seven**
27 **options result in competition for land (*medium confidence*).** A large number of response options do
28 not require dedicated land, including several land management options, all value chain options, and all
29 risk management options. Four options could greatly increase competition for land if applied at scale:
30 afforestation, reforestation, and land used to provide feedstock for BECCS and biochar, with three
31 further options: reduced grassland conversion to croplands, restoration and reduced conversion of
32 peatlands and restoration and reduced conversion of coastal wetlands having smaller or variable
33 impacts on competition for land. Other options such as reduced deforestation and degradation, restrict
34 land conversion for other options and uses. Expansion of the current area of managed land into natural
35 ecosystems could have negative consequences for other land challenges, lead to the loss of
36 biodiversity, and adversely affect a range of NCPs (*high confidence*). {6.3.6, 6.4}

37 **Some options, such as bioenergy and BECCS, are scale dependent. The climate change**
38 **mitigation potential for bioenergy and BECCS is large (up to 11 GtCO₂ yr⁻¹); however, the**
39 **effects of bioenergy production on land degradation, food insecurity, water scarcity, GHG**
40 **emissions, and other environmental goals are scale and context specific (*high confidence*).** These
41 effects depend on the scale of deployment, initial land use, land type, bioenergy feedstock, initial
42 carbon stocks, climatic region and management regime (*high confidence*). Large areas of monoculture
43 bioenergy crops that displace other land uses can result in land competition, with adverse effects for
44 food production, food consumption, and thus food security, as well as adverse effects for land
45 degradation, biodiversity, and water scarcity (*medium confidence*). However, integration of bioenergy
46 into sustainably managed agricultural landscapes can ameliorate these challenges (*medium*
47 *confidence*). {6.2, 6.3, 6.4, Cross-Chapter Box 7 in this Chapter}

1 **Response options are interlinked; some options (e.g., land sparing and sustainable land**
2 **management options) can enhance the co-benefits or increase the potential for other options**
3 **(medium confidence).** Some response options can be more effective when applied together (*medium*
4 *confidence*); for example, dietary change and waste reduction expand the potential to apply other
5 options by freeing as much as 5.8 Mkm² (0.8-2.4 Mkm² for dietary change; ~2 Mkm² for reduced
6 post-harvest losses, and 1.4 Mkm² for reduced food waste) of land (*low confidence*). Integrated water
7 management and increased soil organic carbon can increase food productivity in some circumstances.
8 {6.4}

9 **Other response options (e.g., options that require land) may conflict; as a result, the potentials**
10 **for response options are not all additive, and a total potential from the land is currently**
11 **unknown (high confidence).** Combining some sets of options (e.g., those that compete for land) may
12 mean that maximum potentials cannot be realised, for example reforestation, afforestation, and
13 bioenergy and BECCS all compete for the same finite land resource so the combined potential is
14 much lower than the sum of potentials of each individual option calculated in the absence of
15 alternative uses of the land (*high confidence*). Given the interlinkages among response options and
16 that mitigation potentials for individual options assume that they are applied to all suitable land, the
17 total mitigation potential is much lower than the sum of the mitigation potential of the individual
18 response options (*high confidence*). {6.4}

19 **The feasibility of response options, including those with multiple co-benefits, is limited due to**
20 **economic, technological, institutional, socio-cultural, environmental and geophysical barriers**
21 **(high confidence).** A number of response options (e.g., most agriculture-based land management
22 options, forest management, reforestation and restoration) have already been implemented widely to
23 date (*high confidence*). There is robust evidence that many other response options can deliver co-
24 benefits across the range of land challenges, yet these are not being implemented. This limited
25 application is evidence that multiple barriers to implementation of response options exist (*high*
26 *confidence*). {6.3, 6.4}

27 **Coordinated action is required across a range of actors, including business, consumers, land**
28 **managers, indigenous and local communities and policymakers to create enabling conditions for**
29 **adoption of response options (high confidence).** The response options assessed face a variety of
30 barriers to implementation (economic, technological, institutional, socio-cultural, environmental and
31 geophysical) that require action across multiple actors to overcome (*high confidence*). There are a
32 variety of response options available at different scales that could form portfolios of measures applied
33 by different stakeholders from farm to international scales. For example, agricultural diversification
34 and use of local seeds by smallholders can be particularly useful poverty reduction and biodiversity
35 conservation measures, but are only successful when higher scales, such as national and international
36 markets and supply-chains, also value these goods in trade regimes, and consumers see the benefits of
37 purchasing these goods. However, the land and food sectors face particular challenges of institutional
38 fragmentation, and often suffer from a lack of engagement between stakeholders at different scales
39 (*medium confidence*). {6.3, 6.4}

40 **Delayed action will result in an increased need for response to land challenges and a decreased**
41 **potential for land-based response options due to climate change and other pressures (high**
42 **confidence).** For example, failure to mitigate climate change will increase requirements for adaptation
43 and may reduce the efficacy of future land-based mitigation options (*high confidence*). The potential
44 for some land management options decreases as climate change increases; for example, climate alters
45 the sink capacity for soil and vegetation carbon sequestration, reducing the potential for increased soil
46 organic carbon (*high confidence*). Other options (e.g., reduced deforestation and degradation) prevent
47 further detrimental effects to the land surface; delaying these options could lead to increased
48 deforestation, conversion, or degradation, serving as increased sources of GHGs and having

1 concomitant negative impacts on NCPs (*medium confidence*). Carbon dioxide removal (CDR)
2 options, like reforestation, afforestation, bioenergy and BECCS, are used to compensate for
3 unavoidable emissions in other sectors; delayed action will result in larger and more rapid deployment
4 later (*high confidence*). Some response options will not be possible if action is delayed too long; for
5 example, peatland restoration might not be possible after certain thresholds of degradation have been
6 exceeded, meaning that peatlands could not be restored in certain locations (*medium confidence*) {6.2,
7 6.3, 6.4}.

8 **Early action, however, has challenges including technological readiness, upscaling, and**
9 **institutional barriers (*high confidence*).** Some of the response options have technological barriers
10 that may limit their wide-scale application in the near-term (*high confidence*). Some response options,
11 e.g., BECCS, have only been implemented at small-scale demonstration facilities; challenges exist
12 with upscaling these options to the levels discussed in this Chapter (*medium confidence*). Economic
13 and institutional barriers, including governance, financial incentives and financial resources, limit the
14 near-term adoption of many response options, and ‘policy lags’, by which implementation is delayed
15 by the slowness of the policy implementation cycle, are significant across many options (*medium*
16 *confidence*). Even some actions that initially seemed like ‘easy wins’ have been challenging to
17 implement, with stalled policies for REDD+ providing clear examples of how response options need
18 sufficient funding, institutional support, local buy-in, and clear metrics for success, among other
19 necessary enabling conditions. {6.2, 6.4}

20 **Some response options reduce the consequences of land challenges, but do not address**
21 **underlying drivers (*high confidence*).** For example, management of urban sprawl can help reduce
22 the environmental impact of urban systems; however, such management does not address the
23 socioeconomic and demographic changes driving the expansion of urban areas. By failing to address
24 the underlying drivers, there is a potential for the challenge to re-emerge in the future (*high*
25 *confidence*). {6.4}

26 **Many response options have been practiced in many regions for many years; however, there is**
27 **limited knowledge of the efficacy and broader implications of other response options (*high***
28 ***confidence*).** For the response options with a large evidence base and ample experience, further
29 implementation and upscaling would carry little risk of adverse side-effects (*high confidence*).
30 However, for other options, the risks are larger as the knowledge gaps are greater; for example,
31 uncertainty in the economic and social aspects of many land response options hampers the ability to
32 predict their effects (*medium confidence*). Furthermore, Integrated Assessment Models, like those
33 used to develop the pathways in the IPCC Special Report on Global Warming of 1.5°C (SR15), omit
34 many of these response options and do not assess implications for all land challenges (*high*
35 *confidence*). {6.4}

36

37

1 **6.1 Introduction**

2 **6.1.1 Context of this chapter**

3 This chapter focuses on the interlinkages between response options¹ to deliver climate mitigation and
4 adaptation, to address desertification and land degradation, and to enhance food security, and also
5 assesses reported impacts on Nature's Contributions to People (NCP) and contributions to the UN
6 Sustainable Development Goals (SDG). By identifying which options provide the most co-benefits
7 with the fewest adverse side-effects, this chapter aims to provide *integrated response options* that
8 could co-deliver across the range of challenges. This chapter *does not consider*, in isolation, response
9 options that affect only one of climate mitigation, adaptation, desertification, land degradation, or
10 food security, since these are the subjects of Chapters 2–5; this chapter *considers only* interlinkages
11 between response options and two or more of these challenges in the land sector.

12 Since the aim is to assess and provide guidance on integrated response options, each response option
13 is first described and categorised drawing on previous chapters 2-5 (Section 6.2), and their impact on
14 climate mitigation / adaptation, desertification, land degradation, and food security are quantified
15 (Section 6.3). The feasibility of each response option, respect to costs, barriers, saturation and
16 reversibility is then assessed (Section 6.4.1), before considering their sensitivity to future climate
17 change (Section 6.4.2).

18 The *co-benefits* and *adverse side-effects*² of each integrated response option across the five land
19 challenges, and their impacts on the NCP and the SDG, are then assessed in Section 6.4.3. In section
20 6.4.4, the spatial applicability of these integrated response options is assessed in relation to the
21 location of the challenges with the aim of identifying which options have the greatest potential to co-
22 deliver across the challenges, and the contexts and circumstances in which they do so. Interlinkages
23 among response options and challenges in future scenarios are also assessed in 6.4.4. Finally, section
24 6.4.5 discusses the potential consequences of delayed action.

25 In providing this evidence-based assessment, drawing on the relevant literature, this chapter does not
26 assess the merits of policies to deliver these integrated response options - Chapter 7 assesses the
27 various policy options currently available to deliver these interventions - rather this chapter provides
28 an assessment of the integrated response options and their ability to co-deliver across the multiple
29 challenges addressed in this Special Report.

30 **6.1.2 Framing social challenges and acknowledging enabling factors**

31 In this section we outline the approach used in assessing the evidence for interactions between
32 response options to deliver climate mitigation and adaptation, to prevent desertification and land
33 degradation, and to enhance food security. Overall, while defining and presenting the response
34 options to meet these goals is the primary goal of this chapter, we note that these options must not be
35 considered only as technological interventions, or one-off actions. Rather, they need to be understood
36 as responses to socio-ecological challenges whose success will largely depend on external enabling
37 factors. There have been many previous efforts at compiling positive response options that meet
38 numerous Sustainable Development Goals, but which have not resulted in major shifts in

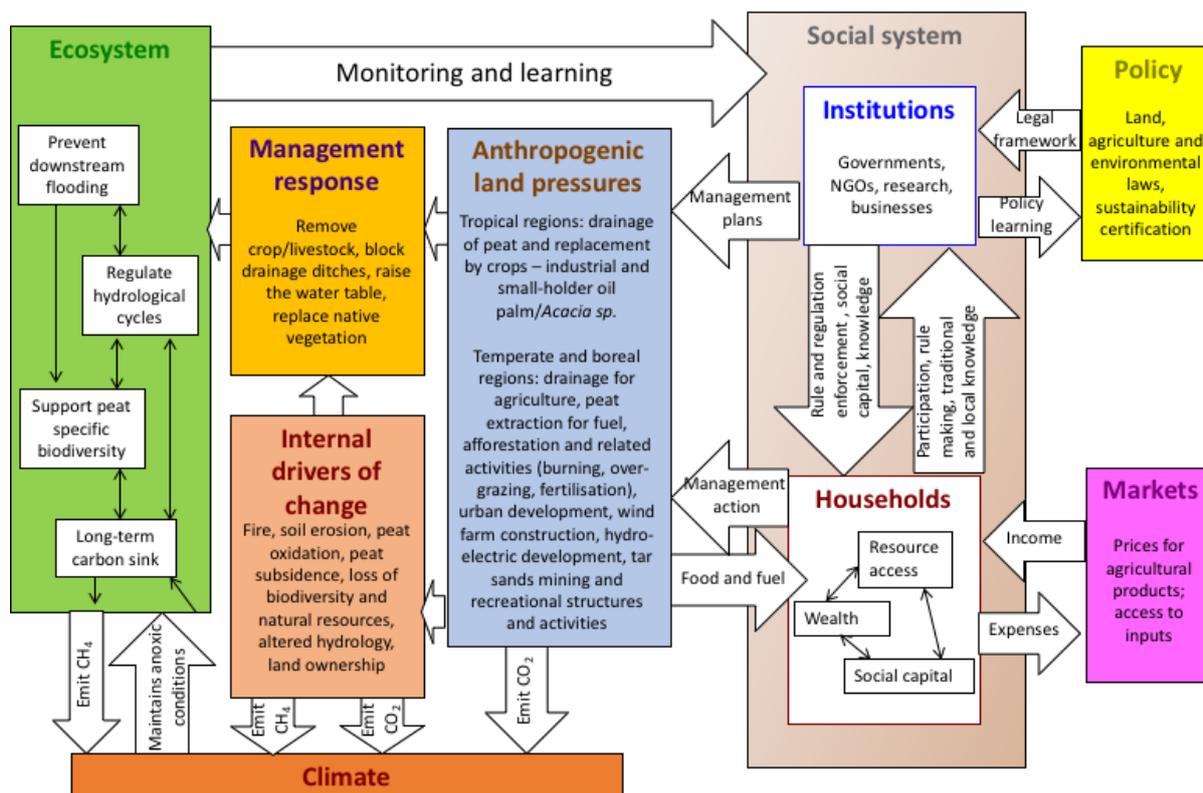
¹ FOOTNOTE: Many of the response options considered are *sustainable land management* options, but several response options are not based on land management, for example those based on value chain management and governance and risk management options

² FOOTNOTE: We use the IPCC AR5 WGIII definitions of co-benefits and adverse side-effect – see glossary. Co-benefits and adverse side-effects can be biophysical and/or socio-economic in nature, and all are assessed as far as the literature allows.

1 implementation; for example, online databases of multiple response options for sustainable land
2 management (SLM), adaptation, and other objectives have been compiled by many donor agencies,
3 including World Overview of Conservation Approaches and Technologies (WOCAT), Climate Adapt,
4 and the Adaptation Knowledge Portal, (Schwilch et al.,2012)³. Yet clearly barriers to adoption
5 remain, or these actions would have been more widely used by now. Much of the scientific literature
6 on barriers to implementing response options focuses on the individual and household level, and
7 discusses limits to adoption, often primarily identified as economic factors (Nigussie et al. 2017;
8 Dallimer et al. 2018). While a useful approach, such studies often are unable to account for the larger
9 enabling factors that might assist in more wide scale implementation (chapter 7 discusses these
10 governance factors and barriers to be overcome in more detail).

11 Instead, this chapter proposes that each response option identified and assessed needs to be
12 understood as an intervention within complex socio-ecological systems (SES) (introduced in Chapter
13 1). In this understanding, physical changes affect human decision-making over land and risk
14 management options, as do economics, policies, and cultural factors, which in turn may drive
15 additional ecological change (Rawlins et al.,2010). This co-evolution of responses within an SES
16 provides a more nuanced understanding of the dynamics between drivers of change and impacts of
17 interventions. Thus, in discussions of the 40 specific response options in this chapter, it must be kept
18 in mind that all need to be contextualised within the specific SES in which they are deployed (see
19 Figure 6.1). Framing response options within SESs also recognises the interactions *between* different
20 response options. However, a major problem within SESs is that the choice and use of different
21 response options requires knowledge of the problems they are aimed at solving, which may be
22 unclear, contested, or not shared equally among stakeholders (Carmenta et al., 2017). Drivers of
23 environmental change often have primarily social or economic rather than technological roots, which
24 requires acknowledgement that response options that do not aim at reducing the drivers of change
25 may thus be less successful (Schwilch et al., 2014).

³ FOOTNOTE: E.g. see <https://qcat.wocat.net/en/wocat/>; <https://climate-adapt.eea.europa.eu>;
<https://www4.unfccc.int/sites/NWPStaging/Pages/Home.aspx>



1
2 **Figure 6.1 Model to represent a social-ecological system of one of the integrated response options in this**
3 **chapter, using restoration and reduced impact of peatlands as an example. The boxes show systems**
4 **(ecosystem, social system), external and internal drivers of change and the management response – here**
5 **enacting the response option. Unless included in the internal drivers of change box, all other drivers of**
6 **change are external (e.g. climate, policy, markets, anthropogenic land pressures). The arrows represent**
7 **how the systems can influence each other, with key drivers of impact written in the arrow in the direction**
8 **of effect.**

9 Response options must also account for the uneven distribution of impacts among populations of
10 both environmental change and intervention responses to this change. Understanding the integrated
11 response options available in a given context requires an understanding of the specificities of social
12 vulnerability, adaptive capacity, and institutional support to assist communities, households and
13 regions to reach their capabilities and achievement of the SDG and other social and land
14 management goals. Vulnerability reflects how assets are distributed within and among communities,
15 shaped by factors that are not easily overcome with technical solutions, including inequality and
16 marginalisation, poverty, and access to resources (Adger et al. 2004; Hallegate et al 2016).
17 Understanding why some people are vulnerable and what structural factors perpetuate this
18 vulnerability requires attention to both micro and meso scales (Tschakert et al. 2013). These
19 vulnerabilities create barriers to adoption of even low-cost high-return response options, such as soil
20 carbon management, that may seem obviously beneficial to implement (Mutoko et al. 2014;
21 Cavanagh et al. 2017). Thus, assessment of the differentiated vulnerabilities that may prevent
22 response option adoption needs to be considered as part of any package of interventions.

23 Adaptive capacity relates to the ability of institutions or people to modify or change characteristics or
24 behaviour so as to cope better with existing or anticipated external stresses (Moss et al. 2001;
25 Brenkert and Malone 2005; Brooks et al. 2005). Adaptive capacity reflects institutional and policy
26 support networks, and has often been associated at the national level with strong developments in the
27 fields of economics, education, health, and governance and political rights (Smit et al. 2001). Areas

1 with low adaptive capacity, as reflected in low Human Development Index scores, might constrain
2 the ability of communities to implement response options (section 6.4.4.1 and Figure 6.7).

3 Further, while environmental changes like land degradation have obvious social and cultural impacts,
4 as discussed in the preceding chapters, so do response options, and thus careful thought is needed
5 about what impacts are expected and what trade-offs are acceptable. One potential way to assess the
6 impact of response interventions relates to the idea of capabilities, a concept first proposed by
7 economist Amartya Sen (Sen 1992). Understanding capability as the “freedom to achieve well-being”
8 frames a problem as being a matter of facilitating what people aspire to do and be, rather than telling
9 them to achieve a standardised or predetermined outcome (Nussbaum and Sen 1993). Thus a
10 capability approach is generally a more flexible and multi-purpose framework, appropriate to an SES
11 understanding because of its open-ended approach (Bockstael and Berkes 2017). Thus, one question
12 for any decision-maker approaching schematics of response options is to determine which response
13 options lead to increased or decreased capabilities for the stakeholders who are the objects of the
14 interventions, given the context of the SES in which the response option will be implemented.

15 Section 6.4.3 examines some of the capabilities that are reflected in the UN Sustainable Development
16 Goals (SDG), such as gender equity and education, and assesses how each of the 40 response options
17 may affect those goals, either positively or negatively, through a review of the available literature.

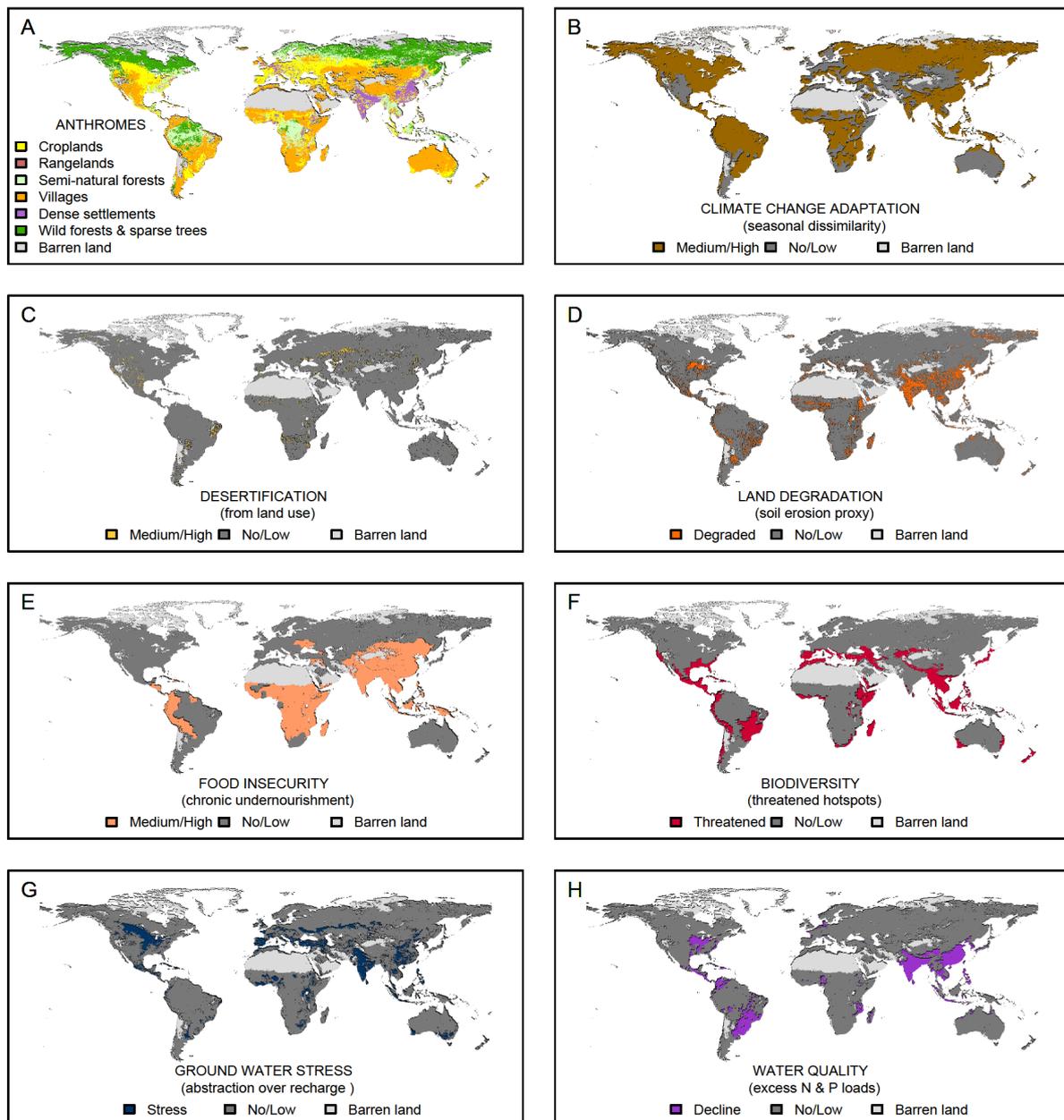
18 **6.1.2.1 Enabling conditions**

19 Response options are not implemented in a vacuum and rely on knowledge production and socio-
20 economic and cultural strategies and approaches embedded within them to be successful. For
21 example, it is well known that “Weak grassroots institutions characterised by low capacity, failure to
22 exploit collective capital and poor knowledge sharing and access to information, are common barriers
23 to sustainable land management and improved food security” (Oloo and Omondi 2017). Achieving
24 broad goals such as reduced poverty or sustainable land management requires conducive enabling
25 conditions, such as attention to gender issues and the involvement of stakeholders, like indigenous
26 peoples, as well as attention to governance, including adaptive governance, stakeholder engagement,
27 and institutional facilitation (see section 6.4.4.3). These enabling conditions – such as gender-
28 sensitive programming or community-based solutions - are not categorised as individual response
29 options in subsequent sections of this chapter because they are conditions that can potentially help
30 improve *all* response options when used in tandem to produce more sustainable outcomes. Chapter 7
31 picks up on these themes and discusses the ways various policies to implement response options have
32 tried to minimise unwanted social and economic impacts on participants in more depth, through
33 deeper analysis of concepts such as citizen science and adaptive governance. Here we simply note the
34 importance of assessing the contexts within which response options will be delivered, as no two
35 situations are the same, and no single response option is likely to be a ‘silver bullet’ to solve all land-
36 climate problems, as each option comes with potential challenges and trade-offs (section 6.2), barriers
37 to implementation (section 6.4.1), interactions with other sectors of society (section 6.4.3), and
38 potential environmental limitations (section 6.4.4).

39 **6.1.3 Challenges and response options in current and historical interventions**

40 Land-based systems are exposed to multiple overlapping challenges including climate change
41 (adaptation and mitigation), desertification (Chapter 3), land degradation (Chapter 4) and food
42 insecurity (Chapter 5), as well as loss of biodiversity, ground water stress (from over-abstraction) and
43 water quality. The spatial distribution of these individual land-based challenges is shown in Figure
44 6.2, based on recent studies and using the following indicators:

- 1 • Desertification attributed to land use is estimated from vegetation remote sensing (Figure
2 3.7c), mean annual change in NDVImax < -0.001 (between 1982-2015) in dryland areas
3 (Aridity Index > 0.65), noting however that desertification has multiple causes (Chapter 3);
4 • Land degradation (see Chapter 4) is based on a soil erosion (Borrelli et al. 2017) proxy
5 (annual erosion rate of 3 t ha⁻¹ or above);
6 • The climate change challenge for adaptation is based on a dissimilarity index of monthly
7 means of temperature and precipitation between current and end of century scenarios
8 (dissimilarity index equal to 0.7 or above, Netzel and Stepinski 2018), noting however that
9 rapid warming could occur in all land regions (Chapter 2);
10 • The food security challenge is estimated as the prevalence of chronic undernourishment
11 (higher or equal to 5%) by country in 2015 (FAO 2017a), noting however that food security
12 has several dimensions (see Chapter 5);
13 • The biodiversity challenge uses threatened terrestrial biodiversity hotspots (areas where
14 exceptional concentrations of endemic species are undergoing exceptional loss of habitat,
15 (Mittermeier et al. 2011), noting however that biodiversity concerns more than just
16 threatened endemic species;
17 • The groundwater stress challenge is estimated as groundwater abstraction over recharge
18 ratios above one (Gassert et al. 2014) in agricultural areas (croplands and villages);
19 • The water quality challenge is estimated as critical loads (higher or equal to 1000 kg N km⁻²
20 or 50 kg P km⁻²) of nitrogen (N) and phosphorus (P) (Xie and Ringler 2017)
- 21 Overlapping land-based challenges affect all land use categories: croplands, rangelands, semi-natural
22 forests, villages, dense settlements, wild forests and sparse trees and barren lands. These land use
23 categories can be defined as anthropogenic biomes, or anthromes, and their global distribution was
24 mapped by Ellis and Ramankutty (2008) (Figure 6.2).



1
2
3
4
5
6
7
8
9
10

Figure 6.2 Global distributions of land use types and individual land-based challenges. A, land use types (or anthromes, after Ellis and Ramankutty 2008); B, climate change adaptation challenge (estimated from dissimilarity between current and end of century climate scenarios, Netzel and Stepinski 2018); C, desertification challenge (after Chapter 3, Figure 3.7c); D, land degradation challenge (estimated from a soil erosion proxy, one indicator of land degradation Borrelli et al. 2017); E, food security challenge (estimated from chronic undernourishment, a component of food security, FAO 2017a); F, biodiversity challenge (estimated from threatened biodiversity hotspots, a component of biodiversity, Mittermeier et al. 2011); G, groundwater stress challenge (estimated from water over-abstraction, Gassert et al. 2014); H, water quality challenge (estimated from critical N and P loads of water systems, Xie and Ringler 2017).

1 **Table 6.1 Global area of land use types (or anthromes) and current percentage area exposure to**
 2 **individual (overlapping) land-based challenges. See Figure 6.2 and text for further details on criteria for**

| Land use type (anthrome ^a) | Anthrome area | Climate change adaptation (dissimilarity index proxy) ^b | Land degradation (soil erosion proxy) ^c | Desertification (ascribed to land use) ^d | Food security (chronic under nourishment) ^e | Biodiversity (threatened hotspot) ^f | Ground water stress (over abstraction) ^g | Water quality (critical N-P loads) ^h |
|--|--|--|--|---|--|--|---|---|
| | <i>% of ice-free land area¹</i> | <i>% anthrome area exposed to an individual challenge</i> | | | | | | |
| Dense settlement | 1 | 76 | 20 | 3 | 30 | 32 | - | 30 |
| Village | 5 | 70 | 49 | 3 | 78 | 28 | 77 | 59 |
| Cropland | 13 | 68 | 21 | 7 | 28 | 27 | 65 | 20 |
| Rangeland | 26 | 46 | 14 | 7 | 43 | 21 | - | 10 |
| Semi-natural forests | 14 | 91 | 17 | 0.7 | - | 21 | - | 7 |
| Wild forests and sparse trees | 17 | 98 | 4 | 0.5 | - | 2 | - | 0.3 |
| Barren | 19 | 53 | 6 | 0.9 | 2 | 4 | - | 0.4 |
| *Organic soils | 4 | 95 | 10 | 2 | 9 | 13 | - | 6 |
| *Coastal wetlands | 0.6 | 74 | 11 | 2 | 24 | 33 | - | 26 |
| All anthromes | 100 | 69 | 13 | 3.2 | 20 | 15 | 12 | 10 |

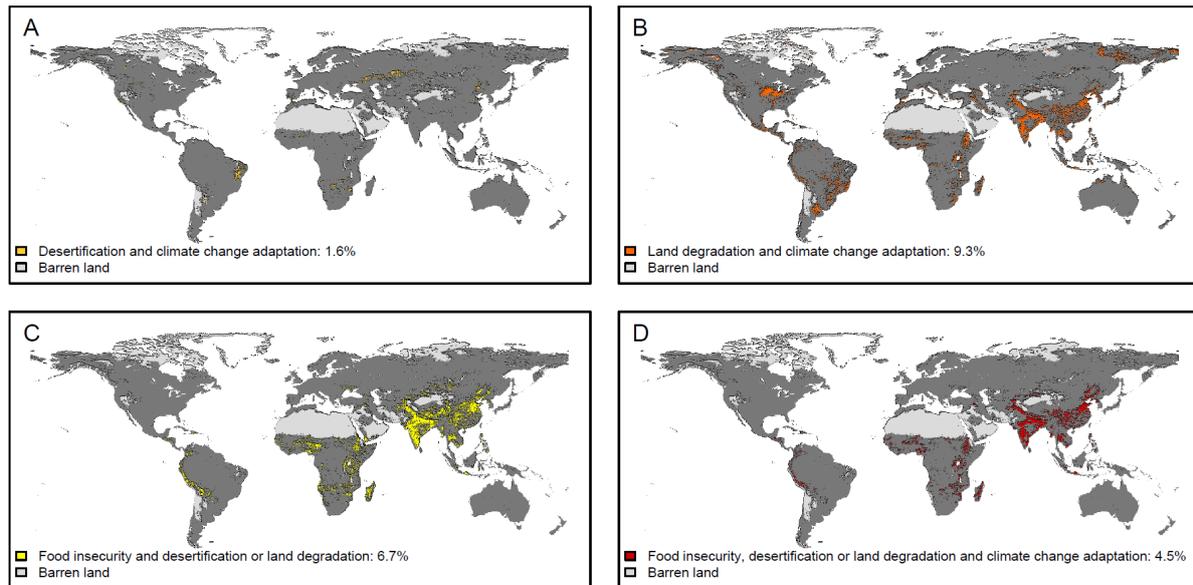
3 **individual challenges.**

4 ^a Ellis and Ramankutty (2008) - the global ice-free land area is estimated at 134 Mkm²; ^b Borrelli et al. 2017; ^c Netzel and
 5 Stepinski 2018; ^d From Figure 3.7c, chapter 3; ^e FAO 2017a; ^f Mittermeier et al. 2011; ^g Gassert et al 2014; ^h Xie and Ringler
 6 2017

7
 8 The majority of the global population is concentrated in dense settlements and villages accounting for
 9 less than 7% of the global ice-free land area, while croplands and rangelands use 39% of land. The
 10 remainder of the ice-free land area (more than half) is used by semi-natural forests, by wild forests
 11 and sparse trees and by barren lands (Table 6.1).

12 Land use types (or anthromes) are exposed to multiple overlapping challenges. Climate change could
 13 induce rapid warming in all land areas (see Chapter 2). In close to 70% of the ice-free land area, the
 14 climate change adaptation challenge could be reinforced by a strong dissimilarity between end of
 15 century and current temperature and precipitation seasonal cycles (Netzel and Stepinski 2018).
 16 Chronic undernourishment (a component of food insecurity) is concentrated in 20% of global ice-free
 17 land area. Severe soil erosion (a proxy of land degradation) and desertification from land use affect 13
 18 and 3% of ice-free land area, respectively. Both groundwater stress and severe water quality decline
 19 (12 and 10% of ice-free land area, respectively) contribute to the water challenge. Threatened
 20 biodiversity hot-spots (15% of ice-free land area) are significant for the biodiversity challenge (Table
 21 6.1).

22 Since land-based challenges overlap, part of the ice-free land area is exposed to combinations of two
 23 or more challenges. For instance, land degradation (severe soil erosion) or desertification from land
 24 use and food insecurity (chronic undernourishment) are combined with a strong climate change
 25 adaptation challenge (dissimilarity in seasonal cycles) in 4.5% of the ice-free land area (Figure 6.3).



1
2 **Figure 6.3 Example of overlap between land challenges. A. Overlap between the desertification (from**
3 **land use) challenge and the climate change adaptation (strong dissimilarity in seasonal cycles) challenge.**
4 **B. Overlap between the land degradation (soil erosion proxy) challenge and the climate change**
5 **adaptation challenge. C. Overlap between the desertification or land degradation challenges and the food**
6 **insecurity (chronic undernourishment) challenge. D. Overlap between challenges shown in C and the**
7 **climate change adaptation challenge. For challenges definitions, see text; references as in Figure 6.2.**

8 The global distribution of land area by the number of overlapping land challenges (Figure 6.4) shows:
9 the least exposure to land challenges in barren lands; less frequent exposure to two or more challenges
10 in wild forests than in semi-natural forests; more frequent exposure to two or more challenges in
11 agricultural anthromes (croplands and rangelands) and dense settlements than in forests; most
12 frequent exposure to 3 or more challenges in villages compared to other land use types. Therefore,
13 land use types intensively used by humans are, on average, exposed to a larger number of challenges
14 than land use types (or anthromes) least exposed to human use.

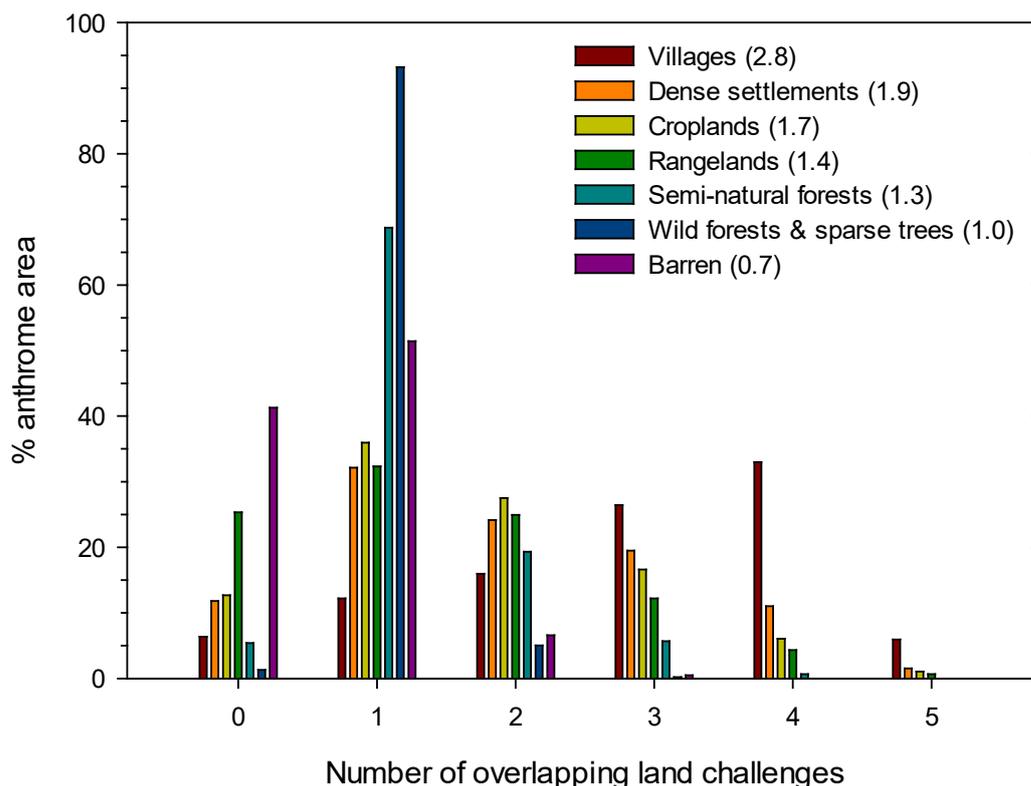


Figure 6.4 Percentage distribution of land use type (or anthrome) area by number of overlapping land challenges for the villages, dense settlements, croplands, rangelands, semi-natural forests, wild forests & sparse trees and barren land use types. Values in brackets show the mean number of land challenges per land use type. Land challenges include desertification (from land use), land degradation (soil erosion proxy), climate change adaptation (seasonal dissimilarity proxy), food security (chronic undernourishment), biodiversity (threatened hot spots), groundwater stress (over abstraction) and water quality (critical N and P loads).

Case studies located in different world regions are presented for each anthrome, in order to provide historical context on the interlinkages between multiple challenges and responses (Box 6.1: A to E). Taken together, these case studies illustrate the large contrast across anthromes in land-based interventions and the way these interventions respond to combinations of challenges.

Box 6.1 Case studies by anthrome type showing historical interlinkages between land-based challenges and the development of local responses

A. Croplands. Land degradation, groundwater stress and food insecurity: soil and water conservation measures in the Tigray region of Ethiopia

In northern Ethiopia, the Tigray region is a drought-prone area that has been subjected to severe land degradation (Frankl et al. 2013) and to recurrent drought and famine during 1888–1892, 1973–1974 and 1984–1985 (Gebremeskel et al. 2018). The prevalence of stunting and being underweight among children under five years is high (Busse et al. 2017) and the region was again exposed to a severe drought during the strong El Niño event of 2015–2016. Croplands are the dominant land-use type, with approximately 90% of the households relying on small-scale plough-based cultivation. Gullies affect nearly all slopes and frequently exceed 2 m in depth and 5 m in top width. Landsat imagery shows that cropland area peaked in 1984–1986 and increased erosion rates in the 1980s and 1990s caused the drainage density and volume to peak in 1994 (Frankl et al. 2013). Since ca. 2000, the large-scale implementation of Soil and Water Conservation (SWC) measures, integrated catchment management, conservation agriculture and indigenous tree regeneration has started to yield positive effects on the vegetation cover and led to the stabilisation of about 25% of the gullies by 2010 (Frankl

1 et al. 2013). Since 1991, farmers have provided labour for SWC in January as a free service for 20
2 consecutive working days, followed by food for work for the remaining days of the dry season. Most
3 of the degraded landscapes have been restored, with positive impacts over the last two decades on soil
4 fertility, water availability and crop productivity. However, misuse of fertilisers, low survival of tree
5 seedlings and lack of income from exclosures may affect the sustainability of these land restoration
6 measures (Gebremeskel et al. 2018).

7 **B. Rangelands. Biodiversity hotspot, land degradation and climate change: pasture** 8 **intensification in the Cerrados of Brazil**

10 Cerrados are a tropical savannah ecoregion in Brazil corresponding to a biodiversity hot spot with less
11 than 2% of its region protected in national parks and conservation areas (Cava et al. 2018). Extensive
12 cattle ranching (limited mechanisation, low use of fertiliser and seed inputs) has led to pasture
13 expansion, including clearing forests to secure properties rights, occurring mainly over 1950–1975
14 (Martha et al. 2012). Despite observed productivity gains made over the last three decades (Martha et
15 al. 2012), more than half of the pasture area is degraded to some extent and challenges remain to
16 reverse grassland degradation while accommodating growing demand and simultaneously avoiding
17 the conversion of natural habitats (de Oliveira Silva et al. 2018). The largest share of production is on
18 unfertilised pastures, often sown with perennial forage grasses of African origin, mainly *Brachiaria*
19 spp. (Cardoso et al. 2016). This initial intensification era was partly at the expense of significant
20 uncontrolled deforestation and average animal stocking rates remained well below the potential
21 carrying capacity (Strassburg et al. 2014). Changes in land use are difficult to reverse since pasture
22 abandonment does not lead to the spontaneous restoration of old-growth savannah (Cava et al. 2018);
23 moreover pasture to crop conversion is frequent, supporting close to half of cropland expansion in
24 Mato Grosso state over 2000–2013 (Cohn et al. 2016). Pasture intensification through liming,
25 fertilisation and controlled grazing could increase soil organic carbon and reduce net GHG emission
26 intensity per unit meat product, but only at increased investment cost per unit of area (de Oliveira
27 Silva et al. 2017). Scenarios projecting a decoupling between deforestation and increased pasture
28 intensification, provide the basis for a Nationally Determined Contribution (NDC) of Brazil that is
29 potentially consistent with accommodating an upward trend in livestock production to meet increasing
30 demand (de Oliveira Silva et al. 2018). Deforestation in Brazil has declined significantly between
31 2004 and 2014 in the national inventory but recent data and analyses suggest that the decrease in
32 deforestation and the resulting GHG emissions reductions have slowed down or even stopped (UNEP
33 2017).

35 **C. Semi-natural forests. Biodiversity hotspot, land degradation, climate change and food** 36 **insecurity: restoration and resilience of tropical forests in Indonesia**

37 During the last two decades, forest cover in Indonesia declined by 150,000 km² in the period 1990-
38 2000 (Stibig et al. 2014) and approximately 158,000 km² in the period 2000–2012 (Hansen et al.
39 2013a), most of which was converted to agricultural lands (e.g., oil palm, pulpwood plantations).
40 According to recent estimates, deforestation in Indonesia mainly concerns primary forests, including
41 intact and degraded forests, thus leading to biodiversity loss and reduced carbon sequestration
42 potentials (e.g., Margono et al. 2014). For example, Graham et al. (2017) estimated that the following
43 strategies to reduce deforestation and degradation may cost-effectively increase carbon sequestration
44 and reduce carbon emissions in 30 years: reforestation (3.54 Gt CO₂), limiting the expansion of oil
45 palm and timber plantations into forest (3.07 Gt CO₂ and 3.05 Gt CO₂, respectively), reducing illegal
46 logging (2.34 Gt CO₂), and halting illegal forest loss in Protected Areas (1.52 Gt CO₂) at a total cost
47 of 15.7 USD tC⁻¹. The importance of forest mitigation in Indonesia is indicated by the NDC, where
48 between half and two-thirds of the 2030 emission target relative to business-as-usual scenario is from
49 reducing deforestation, forest degradation, peatland drainage and fires (Grassi et al. 2017). Avoiding
50 deforestation and reforestation could have multiple co-benefits by improving biodiversity
51 conservation and employment opportunities, while reducing illegal logging in protected areas.
52 However, these options could also have adverse side-effects if they deprive local communities of
53 access to natural resources (Graham et al. 2017). The adoption of the Roundtable on Sustainable Palm
54 Oil certification in oil palm plantations reduced deforestation rates by approximately 33% in the

1 period 2001–2015 (co-benefits with mitigation), and fire rates much more than for non-certified
2 plantations (Carlson et al. 2018). However, given that large-scale oil palm plantations are one of
3 largest drivers of deforestation in Indonesia, objective information on the baseline trajectory for land
4 clearance for oil palm is needed to further assess commitments, regulations and transparency in
5 plantation development (Gaveau et al. 2016). For adaptation options, the community forestry scheme
6 “Hutan Desa” (Village Forest) in Sumatra and Kalimantan helped to avoid deforestation (co-benefits
7 with mitigation) by between 0.6 and 0.9 ha km⁻² in Sumatra and 0.6 and 0.8 ha km⁻² in Kalimantan in
8 the period 2012–2016; Santika et al. 2017), improve local livelihood options, and restore degraded
9 ecosystems (positive side-effects for NCP provision) (e.g., Pohnan et al. 2015). Finally, the
10 establishment of Ecosystem Restoration Concessions in Indonesia (covering more than 5.5 thousand
11 km² of forests now, and 16 thousand km² allocated for the future) facilitates the planting of
12 commercial timber species (co-benefits with mitigation), while assisting natural regeneration,
13 preserving important habitats and species, and improving local well-being and incomes (positive side-
14 effects for Nature’s Contributions to People provision), at relatively lower costs compared with timber
15 concessions (Silalahi et al. 2017).

16 17 **D. Villages. Land degradation, groundwater overuse, climate change and food insecurity: 18 climate smart villages in India**

19 Indian agriculture, which includes both monsoon-dependent rainfed (58%) and irrigated agriculture, is
20 exposed to climate variability and change. Over the past years, the frequency of droughts, cyclones,
21 and hailstorms has increased, with severe droughts in 8 of 15 years between 2002 and 2017 (Srinivasa
22 Rao et al. 2016; Mujumdar et al. 2017). Such droughts result in large yield declines for major crops
23 like wheat in the Indo-Gangetic plain (Zhang et al. 2017). The development of a submersible pump
24 technology in the 1990s, combined with public policies that provide farmers free electricity for
25 groundwater irrigation, resulted in a dramatic increase in irrigated agriculture (Shah et al. 2012). This
26 shift has led to increased dependence on irrigation from groundwater and induced a groundwater
27 crisis, with large impacts on socio-ecosystems. An increasing number of farmers report bore-well
28 failures either due to excessive pumping of an existing well or a lack of water in new wells. The
29 decrease in the groundwater table level has suppressed the recharge of river beds, turning permanent
30 rivers into ephemeral streams (Srinivasan et al. 2015). Wells have recently been drilled in upland
31 areas, where groundwater irrigation is also increasing (Robert et al. 2017). Additional challenges are
32 declining soil organic matter and fertility under monocultures and rice/wheat systems. Unoccupied
33 land is scarce, meaning that the potential for expanding the area farmed is very limited (Aggarwal et
34 al. 2018). In rural areas, diets are deficient in protein, dietary fibre and iron, and mainly comprised of
35 cereals and pulses grown and/or procured through welfare programs (Vatsala et al. 2017). Cultivators
36 are often indebted and suicide rates are much higher than the national average, especially for those
37 strongly indebted (Merriott 2016). Widespread use of diesel pumps for irrigation, especially for
38 paddies, high use of inorganic fertilisers and crop residue burning lead to high GHG emissions
39 (Aggarwal et al. 2018). The Climate-Smart Village (CSV) approach aims at increasing farm yield,
40 income, input use efficiency (water, nutrients, and energy) and reducing GHG emissions (Aggarwal et
41 al. 2018). Climate-smart agriculture interventions are considered in a broad sense by including
42 practices, technologies, climate information services, insurance, institutions, policies, and finance.
43 Options differ based on the CSV site, its agro-ecological characteristics, level of development, and the
44 capacity and interest of farmers and the local government (Aggarwal et al. 2018). Selected
45 interventions included crop diversification, conservation agriculture (minimum tillage, residue
46 retention, laser levelling), improved varieties, weather-based insurance, agro-advisory services,
47 precision agriculture and agroforestry. Farmers’ cooperatives were established to hire farm
48 machinery, secure government credit for inputs, and share experiences and knowledge. Tillage
49 practices and residue incorporation increased rice–wheat yields by 5–37%, increased income by 28–
50 40%, reduced GHG emissions by 16–25%, and increased water-use efficiency by 30% (Jat et al.,
51 2014). The resulting portfolio of options proposed by the CSV approach has been integrated with the
52 agricultural development strategy of some states like Haryana.

53 54 **E. Dense settlements. Climate change and food: green infrastructures**

1 Extreme heat events have led to particularly high rates of mortality and morbidity in cities as urban
2 populations are pushed beyond their adaptive capacities, leading to an increase in mortality rates of
3 30–130% in major cities in developed countries (Norton et al. 2015). Increased mortality and
4 morbidity from extreme heat events are exacerbated in urban populations by the urban heat island
5 effect (Gabriel and Endlicher 2011; Schatz and Kucharik 2015), which can be limited by developing
6 green infrastructure in cities. Urban green infrastructure includes public and private green spaces,
7 including remnant native vegetation, parks, private gardens, golf courses, street trees, urban farming
8 and more engineered options such as green roofs, green walls, biofilters and raingardens (Norton et al.
9 2015). Increasing the amount of vegetation, or green infrastructure, in a city is one way to help reduce
10 urban air temperature maxima and variation. Increasing vegetation by 10% in Melbourne, Australia
11 was estimated to reduce daytime urban surface temperatures by approximately 1°C during extreme
12 heat events (Coutts and Harris 2013). Urban farming (a type of urban green infrastructure) is largely
13 driven by the desire to reconnect food production and consumption (Whittinghill and Rowe 2012; see
14 Chapter 5). Even though urban farming can only meet a very small share of the overall urban food
15 demand, it provides fresh and local food, especially perishable fruits and crops that are usually
16 shipped from far and sold at high prices (Thomaier et al. 2015). Food-producing urban gardens and
17 farms are often started by grassroots initiatives (Ercilla-Montserrat et al. 2019) that occupy vacant
18 urban spaces. In recent years, a growing number of urban farming projects (termed Zero-Acreage
19 farming, or Z-farming, Thomaier et al. 2015) were established in and on existing buildings, using
20 rooftop spaces or abandoned buildings through contracts between food businesses and building
21 owners. Almost all Z-farms are located in cities with more than 150,000 inhabitants, with a majority
22 in North American cities such as New York City, Chicago and Toronto (Thomaier et al. 2015). They
23 depend on the availability of vacant buildings and roof tops thereby competing with other uses, such
24 as roof-based solar systems. Urban farming, however, has potentially high levels of soil pollution and
25 air pollutants, which may lead to crop contamination and health risks. These adverse effects could be
26 reduced on rooftops (Harada et al. 2019) or in controlled environments.

28 **6.1.4 Challenges represented in future scenarios**

29 In this section, the evolution of several challenges (climate change, mitigation, adaptation,
30 desertification, land degradation, food insecurity, biodiversity and water) in the future are assessed,
31 focusing on global analyses. The effect of response options on these land challenges in the future is
32 discussed in Section 6.4.4. Where possible, studies quantifying these challenges in the Shared Socio-
33 economic Pathways (SSPs) (Chapter 1; Cross-Chapter Box 1: Scenarios, Chapter 1; Cross-Chapter
34 Box 9: Illustrative Climate and Land Pathways, in this chapter; O'Neill et al. 2014), as these studies
35 can be used to assess which future scenarios could experience multiple challenges in the future.

36 *Climate change:* Absent any additional efforts to mitigate, global mean temperature rise is expected to
37 increase by anywhere from 2°C to 7.8°C in 2100 relative to the 1850-1900 reference period (Clarke et
38 al. 2014a; Chapter 2). The level of warming varies depending on the climate model (Collins et al.
39 2013), uncertainties in the Earth system (Clarke et al. 2014), and socioeconomic/technological
40 assumptions (Clarke et al. 2014a; Riahi et al. 2017) Warming over land is 1.2 to 1.4 times higher than
41 global mean temperature rise; warming in the arctic region is 2.4 to 2.6 times higher than warming in
42 the tropics (Collins et al. 2013). Increases in global mean temperature are accompanied by increases
43 in global precipitation; however, the effect varies across regions with some regions projected to see
44 increases in precipitation and others to see decreases (Collins et al. 2013; Chapter 2). Additionally,
45 climate change also has implications for extreme events (e.g., drought, heat waves, etc.), freshwater
46 availability, and other aspects of the terrestrial system (Chapter 2).

47 *Mitigation:* Challenges to mitigation depend on the underlying emissions and “mitigative capacity”,
48 including technology availability, policy institutions, and financial resources (O'Neill et al. 2014b).
49 Challenges to mitigation are high in the SSP3 and SSP5, medium in SSP2, and low in SSP1 and SSP4
50 (O'Neill et al. 2014b, 2017; Riahi et al. 2017a).

1 *Adaptation:* Challenges to adaptation depend on climate risk and adaptive capacity, including
 2 technology availability, effectiveness of institutions, and financial resources (O'Neill et al. 2014b).
 3 Challenges to adaptation are high in the SSP3 and SSP4, medium in SSP2, and low in SSP1 and SSP5
 4 (O'Neill et al. 2014b, 2017; Riahi et al. 2017a).

5 *Desertification:* The combination of climate and land use changes can lead to decreases in soil cover
 6 in drylands (Chapter 3). Population living in drylands is expected to increase by 43% in the SSP2-
 7 Baseline, due to both population increases and an expansion of dryland area (UNCCD 2017).

8 *Land degradation:* Future changes in land use and climate have implications for land degradation,
 9 including impacts on soil erosion, vegetation, fire, and coastal erosion (Chapter 4; Scholes et al.
 10 2018). For example, soil organic carbon is expected to decline by 99 GtCO_{2e} in 2050 in an SSP2-
 11 Baseline scenario, due to both land management and expansion in agricultural area (Brink et al.
 12 2018).

13 *Food insecurity:* Food insecurity in future scenarios varies significantly, depending on socio-
 14 economic development and study. For example, the population at risk of hunger ranges from 0 to 800
 15 million in 2050 (Hasegawa et al. 2015a; Ringler et al. 2016; Fujimori et al. 2018b; Hasegawa et al.
 16 2018; Fujimori et al. 2018a; Baldos and Hertel 2015) and 0–600 million in 2100 (Hasegawa et al.
 17 2015a). Food prices in 2100 in non-mitigation scenarios range from 0.9 to about 2 times their 2005
 18 values (Hasegawa et al. 2015a; Calvin et al. 2014a; Popp et al. 2017). Food insecurity depends on
 19 both income and food prices (Fujimori et al. 2018b). Higher income (e.g., SSP1, SSP5), higher yields
 20 (e.g., SSP1, SSP5), and less meat intensive diets (e.g., SSP1) tend to result in reduced food insecurity
 21 (Hasegawa et al. 2018; Fujimori et al. 2018b).

22 *Biodiversity:* Future species extinction rates vary from modest declines to 100-fold increases from
 23 20th century rates, depending on the species (e.g., plants, vertebrates, invertebrates, birds, fish,
 24 corals), the degree of land-use change, the level of climate change, and assumptions about migration
 25 (Pereira et al., 2010). Mean species abundance (MSA) is also estimated to decline in the future by 10–
 26 20% in 2050 (Vuuren et al., 2015; Pereira et al. 2010). Scenarios with greater cropland expansion lead
 27 to larger declines in MSA (UNCCD 2017) and species richness (Newbold et al., 2015).

28 *Water stress:* Changes in both water supply (due to climate change) and water demand (due to
 29 socioeconomic development) in the future have implications for water stress. Water withdrawals for
 30 irrigation increase from about 2500 km³ yr⁻¹ in 2005 to between 2900 and 9000 km³ yr⁻¹ at the end of
 31 the century (Chaturvedi et al. 2013; Kim et al. 2016; Bonsch et al., 2015; Wada and Bierkens 2014;
 32 Graham et al. 2018; Hejazi et al. 2014); total water withdrawals at the end of the century range from
 33 5000 to 13000 km³ yr⁻¹ (Wada and Bierkens 2014a; Hejazi et al. 2014a; Graham et al. 2018; Kim et
 34 al. 2016). The magnitude of change in both irrigation and total water withdrawals depend on
 35 population, income, and technology (Hejazi et al. 2014a; Graham et al. 2018a). The combined effect
 36 of changes in water supply and water demand will lead to an increase of between 1 and 6 billion
 37 people living in water stressed areas (Schlosser et al. 2014; Hanasaki et al. 2013a; Hejazi et al.
 38 2014c). Changes in water quality are not assessed here but could be important (Liu et al. 2017).

39 *Scenarios with Multiple Challenges:* Table 6.2 summarises the challenges across the five SSP
 40 Baseline scenarios.

41 **Table 6.2: Assessment of future challenges to climate change, mitigation, adaptation, desertification, land**
 42 **degradation, food insecurity, water stress, and biodiversity in the SSP Baseline scenarios**

| SSP | Summary of Challenges |
|------|--|
| SSP1 | The SSP1 (van Vuuren et al. 2017b) has low challenges to mitigation and adaptation. The resulting Baseline scenario includes: <ul style="list-style-type: none"> Continued, but moderate, <i>climate change</i>: global mean temperature increases by 3 to 3.5°C in |

| | |
|------|---|
| | <p>2100 (Riahi et al. 2017a; Huppmann et al. 2018),</p> <ul style="list-style-type: none"> • Low levels of <i>food insecurity</i>: malnourishment is eliminated by 2050 (Hasegawa et al. 2015b), • Declines in <i>biodiversity</i>: biodiversity loss increases from 34% in 2010 to 38% in 2100 (UNCCD 2017), and • High <i>water stress</i>: global water withdrawals decline slightly from the baseline in 2071-2100, but ~2.6 billion people live in water stressed areas (Hanasaki et al. 2013b). <p>Additionally, this scenario is likely to have lower challenges related to desertification, land degradation, and biodiversity loss than the SSP2 as it has lower population, lower land use change and lower climate change (Riahi et al. 2017a).</p> |
| SSP2 | <p>The SSP2 (Fricko et al. 2017) is a scenario with medium challenges to mitigation and high challenges to adaptation. The resulting Baseline scenario includes:</p> <ul style="list-style-type: none"> • Continued <i>climate change</i>: global mean temperature increases by 3.8 to 4.3°C in 2100 (Fricko et al. 2017; Riahi et al. 2017a; Huppmann et al. 2018), • Increased challenges related to <i>desertification</i>: the population living in drylands is expected to increase by 43% in 2050 (UNCCD 2017), • Increased <i>land degradation</i>: soil organic carbon is expected to decline by 99 GtCO_{2e} in 2050 (Brink et al. 2018), • Low levels of <i>food insecurity</i>: malnourishment is eliminated by 2100 (Hasegawa et al. 2015b), • Declines in <i>biodiversity</i>: biodiversity loss increases from 34% in 2010 to 43% in 2100 (UNCCD 2017), and • High <i>water stress</i>: global water withdrawals nearly doubles from the baseline in 2071-2100, with ~4 billion people living in water stressed areas (Hanasaki et al. 2013). |
| SSP3 | <p>The SSP3 (Fujimori et al.,2017) is a scenario with high challenges to mitigation and high challenges to adaptation. The resulting Baseline scenario includes:</p> <ul style="list-style-type: none"> • Continued <i>climate change</i>: global mean temperature increases by 4 to 4.8°C in 2100 (Riahi et al. 2017a; Huppmann et al. 2018), • High levels of <i>food insecurity</i>: about 600 million malnourished in 2100 (Hasegawa et al. 2015b), • Declines in <i>biodiversity</i>: biodiversity loss increases from 34% in 2010 to 46% in 2100 (UNCCD 2017), and • High <i>water stress</i>: global water withdrawals more than double from the baseline in 2071-2100, with ~5.5 billion people living in water stressed areas (Hanasaki et al. 2013). <p>Additionally, this scenario is likely to have higher challenges to desertification, land degradation, and biodiversity loss than the SSP2 as it has higher population, higher land use change and higher climate change (Riahi et al. 2017a).</p> |
| SSP4 | <p>The SSP4 (Calvin et al. 2017a) has high challenges to adaptation but low challenges to mitigation. The resulting Baseline scenario includes:</p> <ul style="list-style-type: none"> • Continued <i>climate change</i>: global mean temperature increases by 3.4 to 3.8°C in 2100 (Calvin et al. 2017b; Riahi et al. 2017a; Huppmann et al. 2018), • High levels of <i>food insecurity</i>: about 400 million malnourished in 2100 (Hasegawa et al. 2015b), and • High <i>water stress</i>: about 3.5 billion people live in water stressed areas in 2100 (Hanasaki et al. 2013). <p>Additionally, this scenario is likely to have similar effects on biodiversity loss as the SSP2 as it has similar land use change and similar climate change (Riahi et al. 2017a).</p> |
| SSP5 | <p>The SSP5 (Kriegler et al. 2017) has high challenges to mitigation but low challenges to adaptation. The resulting Baseline scenario includes:</p> <ul style="list-style-type: none"> • Continued <i>climate change</i>: global mean temperature increases by 4.6 to 5.4°C in 2100 (Kriegler et al. 2017; Riahi et al. 2017a; Huppmann et al. 2018), • Low levels of <i>food insecurity</i>: malnourishment is eliminated by 2050 (Hasegawa et al. 2015b), and • Increased water use and water scarcity: global water withdrawals increase by ~80% in 2071- |

| |
|---|
| <p>2100 with nearly 50% of the population living in water stressed areas (Hanasaki et al. 2013b). Additionally, this scenario is likely to have higher effects on biodiversity loss as the SSP2 as it has similar land use change and higher climate change (Riahi et al. 2017a).</p> |
|---|

6.2 Response options, co-benefits and adverse side-effects across the land challenges

This section describes the integrated response options available to address the land challenges of climate change mitigation, climate change adaptation, desertification, land degradation and food security. These can be categorised into options that rely on a) land management, b) value chain management and c) risk management (Figure 6.5). The land management integrated response options can be grouped according to those that are applied in agriculture, in forests, on soils, in other/all ecosystems and those that are applied specifically for carbon dioxide removal (CDR). The value chain management integrated response options can be categorised as those based demand management and supply management. The risk management options are grouped together (Figure 6.5).

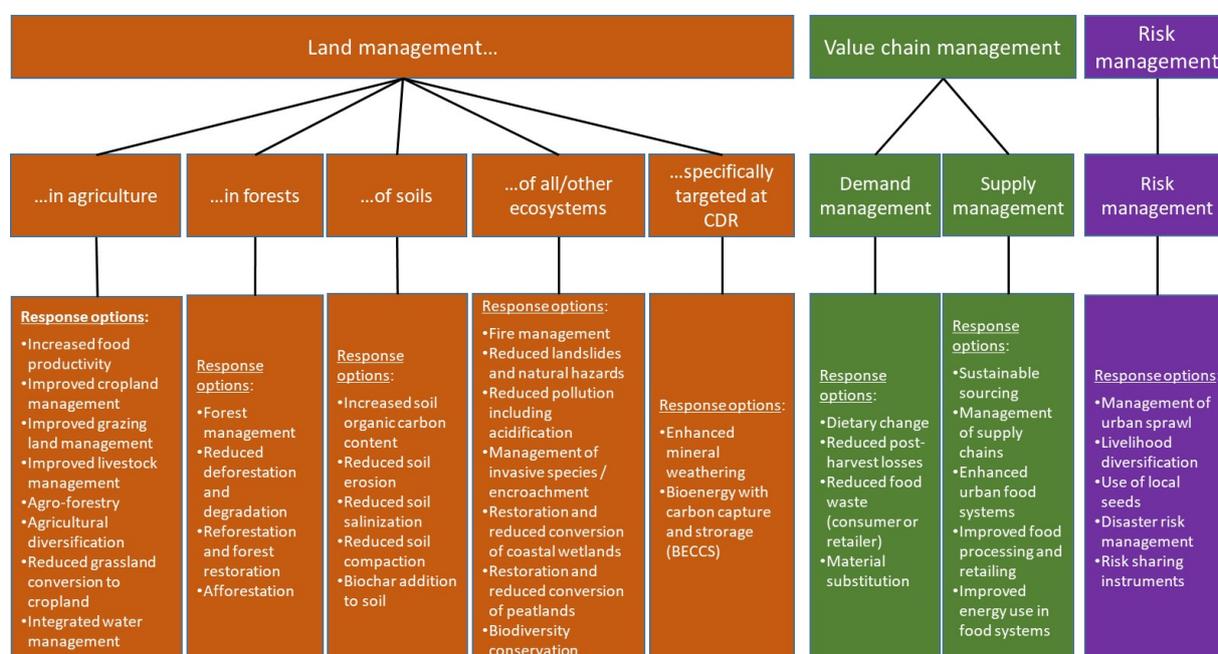


Figure 6.5 Broad categorisation of response options categorised into three main classes and eight sub-classes.

Note that the integrated response options are not mutually exclusive (e.g. cropland management might also increase soil organic matter stocks), and a number of the integrated response options are comprised of a number of practices (e.g., improved cropland management is a collection of practices consisting of a) management of the crop: including high input carbon practices, e. g., improved crop varieties, crop rotation, use of cover crops, perennial cropping systems, agricultural biotechnology; b) nutrient management: including optimised fertiliser application rate, fertiliser type [organic and mineral], timing, precision application, inhibitors; c) reduced tillage intensity and residue retention; d) improved water management: including drainage of waterlogged mineral soils and irrigation of crops in arid / semi-arid conditions; and e) improved rice management: including water management such as mid-season drainage and improved fertilisation and residue management in paddy rice systems).

In this section we deal only with integrated response options, not the policies that are currently / could be implemented to enable their application; that is the subject of Chapter 7. Also note that enabling

1 conditions such as indigenous and local knowledge, gender issues, governance etc. are not categorised
 2 as integrated response options (see Section 6.1.2). Some suggested methods to address land
 3 challenges are better described as *overarching frameworks* than as integrated response options. For
 4 example, *climate smart agriculture* is a collection of integrated response options aimed at delivering
 5 mitigation and adaptation in agriculture, including improved cropland management, grazing land
 6 management and livestock management. Table 6.3 shows how a number of overarching frameworks
 7 are comprised of a range of integrated response options.

8 Similarly, policy goals, such as *land degradation neutrality* (discussed further in Chapter 7), are not
 9 considered as integrated response options. For this reason, *land degradation neutrality*, and
 10 overarching frameworks, such as those described in Table 6.3 do not appear as response options in the
 11 following sections, but the component integrated response options that contribute to these policy goals
 12 or over-arching frameworks are addressed in detail.

13 **Table 6.3 Examples of overarching frameworks that consist of a range of response options, showing how**
 14 **various response options contribute to the overarching frameworks**

| Framework (definition used) | Nature based solutions (IUCN) | Agro-ecology (FAO) | Climate smart agriculture (FAO) | Ecosystem based adaptation (CBD) | Conservation agriculture (FAO) | Community based adaptation (IIED) | Integrated landscape management including integrated coastal zone management (FAO) | Precision agriculture (FAO) | Sustainable forest management (UN) | Sustainable intensification (Cross Chapter Box 5; Chapter 5) | Organic agriculture (FAO) |
|--|-------------------------------|--------------------|---------------------------------|----------------------------------|--------------------------------|-----------------------------------|--|-----------------------------|------------------------------------|--|---------------------------|
| Response options based on land management | | | | | | | | | | | |
| Increased food productivity | | | X | | X | | X | X | | X | |
| Improved cropland management | | X | X | | X | X | X | X | | X | X |
| Improved grazing land management | | X | X | X | | X | X | | | X | X |
| Improved livestock management | | X | X | | | X | X | | | X | X |
| Agroforestry | | X | X | X | | X | X | | | X | X |
| Agricultural diversification | | X | X | | | | X | | | X | X |
| Reduced grassland conversion to cropland | | X | | X | | X | X | | | | |
| Integrated water management | X | X | X | X | X | X | X | X | | X | X |
| Improved forest management | X | | | X | | X | X | | X | | |
| Reduced deforestation and degradation | | X | | X | | X | X | | | | |
| Reforestation and forest restoration | X | X | | X | | X | X | | X | | |
| Afforestation | | | | X | | X | X | | | | |
| Increased soil organic carbon content | | X | X | X | X | | X | | | X | X |

Subject to Copy-editing

| | | | | | | | | | | | |
|--|---|---|---|---|---|---|---|---|---|---|---|
| Reduced soil erosion | | x | x | x | x | | x | | | x | X |
| Reduced soil salinisation | | x | x | x | x | | x | x | | x | X |
| Reduced soil compaction | | x | x | x | x | | x | | | x | X |
| Biochar addition to soil | | x | x | | | | | | | | |
| Fire management | | x | x | x | | x | x | | x | | |
| Reduced landslides and natural hazards | | x | x | x | | x | x | | | | |
| Reduced pollution including acidification | | | | | | | x | x | | x | X |
| Management of invasive species / encroachment | x | x | | x | | x | x | | x | | X |
| Restoration and reduced conversion of coastal wetlands | | x | | x | | x | x | | | | |
| Restoration and reduced conversion of peatlands | | x | x | x | | x | x | | | | |
| Biodiversity conservation | x | x | x | x | x | x | x | | x | x | |
| Enhanced weathering of minerals | | | | | | | | | | | |
| Bioenergy and BECCS | | | | | | | x | | | | |
| <u>Response options based on value chain management</u> | | | | | | | | | | | |
| Dietary change | | x | | | | | | | | | x |
| Reduced post-harvest losses | | x | x | | | x | | x | | | x |
| Reduced food waste (consumer or retailer) | | x | | | | | | | | | |
| Material substitution | | | | | | | | | | | |
| Sustainable sourcing | | x | x | | | x | x | | | | x |
| Management of supply chains | | x | x | | | | | | | | |
| Enhanced urban food systems | | x | x | | | x | x | x | | x | x |
| Improved food processing and retailing | | x | | | | | | | | | |
| Improved energy use in food systems | | x | x | | x | | | x | | x | |
| <u>Response options based on risk management</u> | | | | | | | | | | | |
| Management of urban sprawl | | | | x | | x | x | | | | |
| Livelihood diversification | | x | x | x | | x | x | x | | | |
| Use of local seeds | x | x | x | x | | x | x | | | | |
| Disaster risk management | x | | | x | | x | x | | | | x |
| Risk sharing instruments | | | | | | | | | | x | |

1

2 The SR15 considered a range of response options (from a mitigation / adaptation perspective only).
3 Table 6.4 shows how the SR15 options map on to the response options considered in this report
4 (SRCCL). Note that this report excludes most of the energy-related options from SR15, as well as
5 green infrastructure and sustainable aquaculture.

6

Table 6.4 Mapping of response options considered in this report (SRCCL) and SR15

| SRCCL Response Option or Options | SR15 Response Option or Options |
|---|---|
| Afforestation | Afforestation |
| Reforestation and forest restoration | Reforestation and reduced land degradation and forest restoration |

| | |
|---|--|
| Agricultural diversification | Mixed crop-livestock systems |
| Agroforestry | Agroforestry and silviculture |
| Biochar addition to soil | Biochar |
| Biodiversity conservation | Biodiversity conservation |
| Bioenergy and BECCS | Biomass use for energy production with carbon capture and sequestration (BECCS) (through combustion, gasification, or fermentation) |
| Dietary change | Dietary changes, reducing meat consumption |
| Disaster risk management | Climate services |
| | Community-based adaptation |
| Enhanced urban food systems | Urban and peri-urban agriculture and forestry |
| Enhanced weathering of minerals | Mineralisation of atmospheric CO ₂ through enhanced weathering of rocks |
| Fire management | Fire management and (ecological) pest control |
| Improved forest management | Forest management |
| Improved cropland management | Methane reductions in rice paddies |
| Improved cropland management | Nitrogen pollution reductions, e.g., by fertiliser reduction, increasing nitrogen fertiliser efficiency, sustainable fertilisers |
| | Precision agriculture |
| | Conservation agriculture |
| Improved food processing and retailing | |
| Improved grazing land management | Livestock and grazing management, for example, methane and ammonia reductions in ruminants through feeding management or feed additives, or manure management for local biogas production to replace traditional biomass use |
| Improved livestock management | |
| | Manure management |
| Increased energy efficiency in food systems | |
| Increased food productivity | Increasing agricultural productivity |
| Increased soil organic carbon content | Changing agricultural practices enhancing soil carbon |
| | Soil carbon enhancement, enhancing carbon sequestration in biota and soils, e.g. with plants with high carbon sequestration potential (also AFOLU measure) |
| Integrated water management | Irrigation efficiency |
| Livelihood diversification | |
| Management of invasive species / encroachment | |
| Management of supply chains | |
| Management of urban sprawl | Urban ecosystem services |
| | climate resilient land use |
| Material substitution | Material substitution of fossil CO ₂ with bio-CO ₂ in industrial application (e.g. the beverage industry) |
| | Carbon Capture and Usage – CCU; bioplastics (bio-based materials replacing fossil fuel uses as feedstock in the production of chemicals and polymers), carbon fibre |
| Reduced soil erosion | |
| Reduced soil compaction | |
| Reduced deforestation | Reduced deforestation, forest protection, avoided forest conversion |
| Reduced food waste (consumer or retailer) | Reduction of food waste (incl. reuse of food processing waste for fodder) |
| Reduced grassland conversion to cropland | |
| Reduced landslides and natural hazards | |
| Reduced pollution including acidification | Reduced air pollution |
| Reduced post-harvest losses | |

| | |
|--|--|
| Reduced soil salinisation | |
| Restoration and reduced conversion of coastal wetlands | Managing coastal stress |
| Restoration and reduced conversion of peatlands | Restoration of wetlands (e.g., coastal and peat-land restoration, blue carbon) and wetlands management |
| Risk sharing instruments | Risk sharing |
| Sustainable sourcing | |
| Use of local seeds | |

1 Before providing the quantitative assessment of the impacts of each response option in addressing
 2 mitigation, adaptation, desertification, land degradation and food security in section 6.3, the integrated
 3 response options are described in section 6.2.1 and any context specificities in the effects are noted.

4 **6.2.1 Integrated response options based on land management**

5 **6.2.1.1 Integrated response options based on land management in agriculture**

6 Integrated response options based on land management in agriculture are described in Table 6.5,
 7 which also notes any context specificities in the effects of the response options and provides the
 8 evidence base.

9 **6.2.1.2 Integrated response options based on land management in forests**

10 Integrated response options based on land management in forests are described in Table 6.6, which
 11 also notes any context specificities in the effects of the response options and provides the evidence
 12 base.

13 **6.2.1.3 Integrated response options based on land management of soils**

14 Integrated response options based on land management of soils are described in Table 6.7, which also
 15 notes any context specificities in the effects of the response options and provides the evidence base.

16 **6.2.1.4 Integrated response options based on land management of all/other ecosystems**

17 Integrated response options based on land management in all/other ecosystems are described in Table
 18 6.8, which also notes any context specificities in the effects of the response options and provides the
 19 evidence base.

20 **6.2.1.5 Integrated response options based on land management specifically for carbon dioxide 21 removal (CDR)**

22 Integrated response options based on land management specifically for CDR are described in Table
 23 6.9, which also notes any context specificities in the effects of the response options and provides the
 24 evidence base.

1

Table 6.5 Integrated response options based on land management in agriculture

| Integrated response option | Description | Context and caveats | Supporting evidence |
|------------------------------|--|---|--|
| Increased food productivity | Increased food productivity arises when the output of food commodities increases per unit of input, e.g. per unit of land or water. It can be realised through many other interventions such as improved cropland, grazing land and livestock management. | Many interventions to increase food production, particularly those predicated on very large inputs of agro-chemicals, have a wide range of negative externalities leading to the proposal of sustainable intensification as a mechanism to deliver future increases in productivity that avoid these adverse outcomes. Intensification through additional input of N fertiliser, for example, would result in negative impacts on climate, soil, water and air pollution. Similarly, if implemented in a way that over-exploits the land significant negative impacts would occur, but if achieved through sustainable intensification, and used to spare land, it could reduce the pressure on land. | Cross-Chapter Box 6 on Agricultural Intensification, Chapter 5; Chapter 3; Burney et al. 2010; Foley et al. 2011; Garnett et al. 2013; Godfray et al. 2010; Lal 2016; Lamb et al. 2016; Lobell et al. 2008.; Shcherbak et al. 2014; Smith et al. 2013; Tilman et al. 2014; Scholes et al. 2018; Balmford et al. 2018 |
| Improved cropland management | Improved cropland management is a collection of practices consisting of a) <i>management of the crop</i> : including high input carbon practices, for example, improved crop varieties, crop rotation, use of cover crops, perennial cropping systems, integrated production systems, crop diversification, agricultural biotechnology, b) <i>nutrient management</i> : including optimised fertiliser application rate, fertiliser type (organic manures, compost and mineral), timing, precision application, nitrification inhibitors, c) <i>reduced tillage intensity and residue retention</i> , d) <i>improved water management</i> : including drainage of waterlogged mineral soils and irrigation of crops in arid / semi-arid conditions, e) <i>improved rice management</i> : including water management such as mid-season drainage and improved fertilisation and residue management in paddy rice systems, and f) <i>biochar application</i> . | Improved cropland management can reduce greenhouse gas emissions and create soil carbon sinks, though if poorly implemented, it could increase N ₂ O and CH ₄ emissions from N fertilisers, crop residues and organic amendments. It can improve resilience of food crop production systems to climate change and can be used to tackle desertification and land degradation by improving sustainable land management. It can also contribute to food security by closing crop yield gaps to increase food productivity. | Chapter 4; Chapter 3; Chapter 2; Chapter 5; Bryan et al. 2009; Chen et al. 2010; Labrière et al. 2015; Lal 2011; Poeplau and Don 2015; Porter et al. 2014a; Smith et al. 2014b; Smith 2008; Tilman et al. 2011 |
| Improved grazing land | Improved grazing land management is a collection of practices consisting of a) <i>management of vegetation</i> : including improved grass varieties / sward composition, deep rooting grasses, | Improved grazing land management can increase soil carbon sinks, reduce greenhouse gas emissions, improve the resilience of grazing lands to future | Chapter 2; Chapter 3; Chapter 4; Chapter 5; Section 6.3; Archer et al. 2011; Briske et al. 2015; |

| | | | |
|-------------------------------|--|---|---|
| management | increased productivity, and nutrient management, b) <i>animal management</i> : including appropriate stocking densities fit to carrying capacity, fodder banks, and fodder diversification, and c) <i>fire management</i> : improved use of fire for sustainable grassland management, including fire prevention and improved prescribed burning (see also fire management as a separate response option; Table 6.8). | climate change, help reduce desertification and land degradation by optimising stocking density and reducing overgrazing, and can enhance food security through improved productivity. | Conant et al. 2017; Herrero et al. 2016; Porter et al. 2014a; Schwilch et al. 2014; Smith et al. 2014b; Tighe et al. 2012 |
| Improved livestock management | Improved livestock management is a collection of practices consisting of a) <i>improved feed and dietary additives</i> (e.g., bioactive compounds, fats), used to increase productivity and reduce emissions from enteric fermentation; b) <i>breeding</i> (e.g., breeds with higher productivity or reduced emissions from enteric fermentation), c) <i>herd management</i> , including decreasing neo-natal mortality, improving sanitary conditions, animal health and herd renewal, and diversifying animal species, d) <i>emerging technologies</i> (of which some are not legally authorised in several countries) such as propionate enhancers, nitrate and sulphate supplements, archaea inhibitors and archaeal vaccines, methanotrophs, acetogens, defaunation of the rumen, bacteriophages and probiotics, ionophores / antibiotics; and e) <i>improved manure management</i> , including manipulation of bedding and storage conditions, anaerobic digesters; biofilters, dietary change and additives, soil-applied and animal-fed nitrification inhibitors, urease inhibitors, fertiliser type, rate and timing, manipulation of manure application practices, and grazing management. | Improved livestock management can reduce greenhouse gas emissions, particularly from enteric methane and manure management. It can improve the resilience of livestock production systems to climate change by breeding better adapted livestock. It can help with desertification and land degradation, e.g. through use of more efficient and adapted breeds to allow reduced stocking densities. Improved livestock sector productivity can also increase food production. | Chapter 2; Chapter 3; Chapter 4; Chapter 5; Archer et al. 2011; Herrero et al. 2016; Miao et al. 2015; Porter et al. 2014a; Rojas-Downing et al. 2017; Smith et al. 2008, 2014b; Squires et al. 2005; Tighe et al. 2012 |
| Agroforestry | Agroforestry involves the deliberate planting of trees in croplands and silvo-pastoral systems. | Agroforestry sequesters carbon in vegetation and soils. The use of leguminous trees can enhance biological N fixation and resilience to climate change. Soil improvement and the provision of perennial vegetation can help to address desertification and land degradation. Agroforestry can increase agricultural productivity, with benefits for food security. Additionally, agroforestry can enable payments to farmers for ecosystem services and reduce vulnerability to climate shocks. | Antwi-Agyei et al. 2014; Mbow et al. 2014a; Mutuo et al. 2005; Rosenstock et al. 2014; Sain et al. 2017; Sida et al. 2018; Vignola et al. 2015; Yirdaw et al. 2017 Benjamin et. al. 2018; Guo et al. 2018; Herder et al. 2017; Mosquera-Losada et al. 2018; Nair et al. 2014; Ram et al. 2017; Santiago-Freijanes et. al. 2018; |

| | | | |
|--|--|--|---|
| Agricultural diversification | Agricultural diversification includes a set of agricultural practices and products obtained in the field that aim to improve the resilience of farmers to climate variability and climate change and to economic risks posed by fluctuating market forces. In general, the agricultural system is shifted from one based on low-value agricultural commodities to one that is more diverse, composed of a basket of higher value-added products. | Agricultural diversification is targeted at adaptation but could also deliver a small carbon sink, depending on how it is implemented. It could reduce pressure on land, benefiting desertification, land degradation, food security and household income. However, the potential to achieve household food security is influenced by the market orientation of a household, livestock ownership, non-agricultural employment opportunities, and available land resources. | Birthal et al. 2015; Campbell et al. 2014; Cohn et al. 2017; Lambin and Meyfroidt 2011; Lipper et al. 2014; Massawe et al. 2016; Pellegrini and Tasciotti 2014; Waha et al. 2018 |
| Reduced grassland conversion to cropland | Grasslands can be converted to croplands by ploughing of grassland and seeding with crops. Since croplands have a lower soil carbon content than grasslands and are also more prone to erosion than grasslands, reducing conversion of grassland to croplands will prevent soil carbon losses by oxidation and soil loss through erosion. These processes can be reduced if the rate of grassland conversion to cropland is reduced. | Stabilising soils by retaining grass cover also improves resilience, benefiting adaptation, desertification and land degradation. Since conversion of grassland to cropland usually occurs to remedy food security challenges, food security could be adversely affected, since more land is required to produce human food from livestock products on grassland than from crops on cropland. | Chapter 3; Chapter 4; Chapter 5; Clark and Tilman 2017; Lal 2001a; de Ruiter et al. 2017; Poore & Nemecek, 2018 |
| Integrated water management | Integrated water management is the process of creating holistic strategies to promote integrated, efficient, equitable and sustainable use of water for agroecosystems. It includes a collection of practices including water-use efficiency and irrigation in arid/semi-arid areas, improvement of soil health through increases in soil organic matter content, and improved cropland management, agroforestry and conservation agriculture. Increasing water availability, and reliability of water for agricultural production, can be achieved by using different techniques of water harvesting, storage, and its judicious utilisation through farm ponds, dams, and community tanks in rainfed agriculture areas can benefit adaptation. | These practices can reduce aquifer and surface water depletion, and prevent over extraction, and the management of climate risks. Many technical innovations, e.g., precision water management, can have benefits for both adaptation and mitigation, although trade-offs are possible. Maintaining the same level of yield through use of site-specific water management-based approach could have benefits for both food security and mitigation. | Chapter 3; Chapter 4; Chapter 5; Brindha and Pavelic 2016; Jat et al. 2016; Jiang 2015; Keesstra et al. 2018; Liu et al. 2017; Nejad 2013; Rao et al. 2017; Shaw et al. 2014; Sapkota et al. 2017; Scott et al. 2011; Waldron et al. 2017 |

1

2

1

Table 6.6 Integrated response options based on land management in forests

| Integrated response option | Description | Context and caveats | Supporting evidence |
|---------------------------------------|--|--|--|
| Improved forest management | Improved forest management refers to management interventions in forests for the purpose of climate change mitigation. It includes a wide variety of practices affecting the growth of trees and the biomass removed, including improved regeneration (natural or artificial) and a better schedule, intensity and execution of operations (thinning, selective logging, final cut; reduced impact logging, etc.). Sustainable forest management is the stewardship and use of forests and forest lands in a way, and at a rate, that maintains their biodiversity, productivity, regeneration capacity, vitality and their potential to fulfil, now and in the future, relevant ecological, economic and social functions, at local, national, and global levels, and that does not cause damage to other ecosystems. | Sustainable forest management can enhance the carbon stock in biomass, dead organic matter, and soil – while providing wood-based products to reduce emissions in other sectors through material and energy substitution. A trade-off exists between different management strategies: higher harvest decreases the carbon in the forest biomass in the short term but increases the carbon in wood products and the potential for substitution effects. Sustainable forest management, also through close-to-nature silvicultural techniques, can potentially offer many co-benefits in terms of climate change mitigation, adaptation, biodiversity conservation, microclimatic regulation, soil erosion protection, coastal area protection and water and flood regulation. Forest management strategies aimed at increasing the biomass stock levels may have adverse side-effects, such as decreasing the stand-level structural complexity, biodiversity and resilience to natural disasters. Forest management also affects albedo and evapotranspiration. | Chapter 2; Chapter 4; D’Amato et al. 2011; Dooley and Kartha 2018a; Ellison et al. 2017; Erb et al. 2017; Grassi et al. 2018; Griscom et al. 2017a; Jantz et al. 2014; Kurz et al. 2016; Locatelli 2011; Luysaert et al. 2018; Nabuurs et al. 2017; Naudts et al. 2016; Putz et al. 2012; Seidl et al. 2014; Smith et al. 2014a; Smyth et al. 2014; Stanturf et al. 2015; Forest Europe 2016 Pingoud et al. 2018 |
| Reduced deforestation and degradation | Reduced deforestation and forest degradation includes conservation of existing carbon pools in forest vegetation and soil by controlling the drivers of deforestation (i.e., commercial and subsistence agriculture, mining, urban expansion) and forest degradation (i.e., overharvesting including fuelwood collection, poor harvesting practices, overgrazing, pest outbreaks, and extreme wildfires), also through establishing protected areas, improving law enforcement, forest governance and land tenure, supporting community forest management and introducing forest certification. | Reducing deforestation and degradation is a major strategy to reduce global GHG emissions. The combination of reduced GHG emissions and biophysical effects results in a large climate mitigation effect, with benefits also at local level. Reduced deforestation preserves biodiversity and ecosystem services more efficiently and at lower costs than afforestation/reforestation. Efforts to reduce deforestation and forest degradation may have potential adverse side-effects, for example, reducing availability of land for farming, restricting the rights and access of local people to forest resources (e.g. firewood), or increasing the dependence of local people to insecure external funding. | Chapter 2; Alkama and Cescatti 2016; Baccini et al. 2017; Barlow et al. 2016; Bayrak et al. 2016; Caplow et al. 2011; Curtis et al. 2018; Dooley and Kartha 2018; Griscom et al. 2017a; Hansen et al. 2013b; Hosonuma et al. 2012; Houghton et al. 2015; Lewis et al. 2015; Pelletier et al. 2016; Rey Benayas et al. 2009 |
| Reforestation | Reforestation is the conversion to forest of land that has | Reforestation is similar to afforestation with respect to the co- | Chapter 2; Dooley and |

| | | | |
|-------------------------------|---|--|---|
| <p>and forest restoration</p> | <p>previously contained forests but that has been converted to some other use. Forest restoration refers to practices aimed at regaining ecological integrity in a deforested or degraded forest landscape. As such, it could fall under reforestation if it were re-establishing trees where they have been lost, or under forest management if it were restoring forests where not all trees have been lost. For practical reasons, here forest restoration is treated together with reforestation.</p> | <p>benefits and adverse side-effects among climate change mitigation, adaptation, desertification, land degradation and food security (see row on Afforestation below). Forest restoration can increase terrestrial carbon stocks in deforested or degraded forest landscapes and can offer many co-benefits in terms of increased resilience of forests to climate change, enhanced connectivity between forest areas and conservation of biodiversity hotspots. Forest restoration may threaten livelihoods and local access to land if subsistence agriculture is targeted.</p> | <p>Kartha 2018; Ellison et al. 2017; Locatelli 2011; Locatelli et al. 2015a; Smith et al. 2014b; Stanturf et al. 2015</p> |
| <p>Afforestation</p> | <p>Afforestation is the conversion to forest of land that historically have not contained forests (see also reforestation).</p> | <p>Afforestation increases terrestrial carbon stocks but can also change the physical properties of land surfaces, such as surface albedo and evapotranspiration with implications for local and global climate. In the tropics, enhanced evapotranspiration cools surface temperatures, reinforcing the climate benefits of CO₂ sequestration in trees. At high latitudes and in areas affected by seasonal snow cover, the decrease in surface albedo after afforestation becomes dominant and causes an annual average warming that counteracts carbon benefits. Net biophysical effects on regional climate from afforestation is seasonal and can reduce the frequency of climate extremes, such as heat waves, improving adaptation to climate change and reducing the vulnerability of people and ecosystems. Afforestation helps to address land degradation and desertification, as forests tend to maintain water quality by reducing runoff, trapping sediments and nutrients, and improving groundwater recharge. However, food security could be hampered since an increase in global forest area can increase food prices through land competition. Other adverse side-effects occur when afforestation is based on non-native species, especially with the risks related to the spread of exotic fast-growing tree species. For example, exotic species can upset the balance of evapotranspiration regimes, with negative impacts on water availability, particularly in dry regions.</p> | <p>Chapter 2; Chapter 3; Chapter 4; Chapter 5; Alkama and Cescatti 2016; Arora and Montenegro 2011; Bonan 2008; Boysen et al. 2017; Brundu and Richardson 2016; Cherubini et al. 2017; Ciais et al. 2013; Ellison et al. 2017; Findell et al. 2017; Idris Medugu et al. 2010; Kongsager et al. 2016; Kreidenweis et al. 2016a; Lejeune et al. 2018.; Li et al. 2015; Locatelli et al. 2015a; Perugini et al. 2017; Salvati et al. 2014; Smith et al. 2013, 2014b; Trabucco et al. 2008;</p> |

1

2

1

Table 6.7 Integrated response options based on land management of soils

| Integrated response option | Description | Context and caveats | Supporting evidence |
|---------------------------------------|---|--|---|
| Increased soil organic carbon content | Practices that increase soil organic matter content include a) <i>land use change</i> to an ecosystem with higher equilibrium soil carbon levels (e.g. from cropland to forest), b) <i>management of the vegetation</i> : including high input carbon practices, for example, improved varieties, rotations and cover crops, perennial cropping systems, biotechnology to increase inputs and recalcitrance of below ground carbon, c) <i>nutrient management and organic material input</i> to increase carbon returns to the soil: including optimised fertiliser and organic material application rate, type, timing and precision application, d) <i>reduced tillage intensity and residue retention</i> , and e) <i>improved water management</i> : including irrigation in arid / semi-arid conditions. | Increasing soil carbon stocks removes CO ₂ from the atmosphere and increases the water holding capacity of the soil thereby conferring resilience to climate change and enhancing adaptation capacity. It is a key strategy for addressing both desertification and land degradation. There is some evidence that crop yields and yield stability increase by increased organic matter content, though some studies show equivocal impacts. Some practices to increase soil organic matter stocks vary in their efficacy. For example, the impact of no till farming and conservation agriculture on soil carbon stocks is often positive, but can be neutral or even negative, depending on the amount of crop residues returned to the soil. If soil organic carbon stocks were increased by increasing fertiliser inputs to increase productivity, emissions of nitrous oxide from fertiliser use could offset any climate benefits arising from carbon sinks. Similarly, if any yield penalty is incurred from practices aimed at increasing soil organic carbon stocks (e.g. through extensification), emissions could be increased through indirect land use change, and there could also be adverse side-effects on food security. | Bestelmeyer and Briske 2012; Cheesman et al. 2016; Frank et al. 2017; Gao et al. 2018; Keesstra et al 2016.; Lal 2016, 2006; Lambin and Meyfroidt 2011; de Moraes Sá et al. 2017; Palm et al. 2014; Pan et al. 2009; Paustian et al. 2016; Powlson et al. 2014, 2016, Smith et al. 2013, 2016a, 2014b; Soussana et al. 2019a; Steinbach et al 2006.; VandenBygaart 2016; Hijbeek et al., 2017; Schjønning et al., 2018; |
| Reduced soil erosion | Soil erosion is the removal of soil from the land surface by water, wind or tillage, which occurs worldwide but it is particularly severe in Asia, Latin America and the Caribbean, and the Near East and North Africa. Soil erosion management includes conservation practices (e.g., the use of minimum tillage or zero tillage, crop rotations and cover crops, rational grazing systems), engineering-like practices (e.g., construction of terraces and contour cropping for controlling water erosion), or forest barriers and strip | The fate of eroded soil carbon is uncertain, with some studies indicating a net source of CO ₂ to the atmosphere and others suggesting a net sink. Reduced soil erosion has benefits for adaptation as it reduces vulnerability of soils to loss under climate extremes, increasing resilience to climate change. Some management practices implemented to control erosion, such as increasing ground cover, can reduce the vulnerability of soils to degradation / landslides, and prevention of soil erosion is a key measure used to tackle desertification. Because it protects the capacity of land to produce food, it also contributes positively to food security. | Chapter 3; Chen 2017; Derpsch et al. 2010; FAO and ITPS 2015; FAO 2015; Garbrecht et al. 2015; Jacinthe and Lal 2001; de Moraes Sá et al. 2017; Poeplau and Don 2015; Smith et al. 2001; Stallard 1998; Lal and Moldenhauer 1987; Van Oost et al. 2007; Lugato et al. 2016; Smith et al. 2005; Lal 2001a |

| | | | |
|---------------------------|--|--|--|
| | cultivation for controlling wind erosion. In eroded soils, the advance of erosion gullies and sand dunes can be limited by increasing plant cover, among other practices. | | |
| Reduced soil salinisation | Soil salinisation is a major process of land degradation that decreases soil fertility and affects agricultural production, aquaculture and forestry. It is a significant component of desertification processes in drylands. Practices to reduce soil salinisation include improvement of water management (e.g., water-use efficiency and irrigation/drainage technology in arid/semi-arid areas, surface and groundwater management), improvement of soil health (through increase in soil organic matter content) and improved cropland, grazing land and livestock management, agroforestry and conservation agriculture. | Techniques to prevent and reverse soil salinisation may have small benefits for mitigation by enhancing carbon sinks. These techniques may benefit adaptation and food security by maintaining existing crop systems and closing yield gaps for rainfed crops. These techniques are central to reducing desertification and land degradation, since soil salinisation is a primary driver of both. | Section 3.5; Chapter 4; Chapter 5; Baumhardt et al. 2015; Dagar et al. 2016a; Datta et al. 2000; DERM 2011; Evans and Sadler 2008; He et al. 2015; D'Odorico et al. 2013; Prathapar 1988; Qadir et al. 2013; Rengasamy 2006; Singh 2009; UNCTAD 2011; Wong et al. 2010 |
| Reduced soil compaction | Reduced soil compaction mainly includes agricultural techniques (e.g. crop rotations, control of livestock density) and control of agricultural traffic. | Techniques to reduce soil compaction have variable impacts on GHG emissions but may benefit adaptation by improving soil climatic resilience. Since soil compaction is a driver of both desertification and land degradation, a reduction of soil compaction could benefit both. It could also help close yield gaps in rainfed crops. | Chamen et al. 2015; Epron et al. 2016; ITPS-FAO 2015; Hamza and Anderson 2005; Soane and van Ouwkerk 1994; Tullberg et al. 2018 |
| Biochar addition to soil | The use of biochar, a solid product of the pyrolysis process, as a soil amendment increases the water-holding capacity of soil. It may therefore provide better access to water and nutrients for crops and other vegetation types (so can form part of cropland, grazing land and improved forest management). | The use of biochar increases carbon stocks in the soil. It can enhance yields in the tropics (but less so in temperate regions), thereby benefiting both adaptation and food security. Since it can improve soil water holding capacity and nutrient use efficiency, and can ameliorate heavy metal pollution and other impacts, it can benefit desertification and land degradation. The positive impacts could be tempered by additional pressure on land if large quantities of biomass are required as feedstock for biochar production. | Chapter 2; Chapter 3; Chapter 4; Chapter 5; Jeffery et al. 2017; Smith 2016; Sohi 2012; Woolf et al. 2010 |

1

2

1

Table 6.8 Integrated response options based on land management of all/other ecosystems

| Integrated response option | Description | Context and caveats | Supporting evidence |
|---|---|--|---|
| Fire management | Fire management is a land management option aimed at safeguarding life, property and resources through the prevention, detection, control, restriction and suppression of fire in forest and other vegetation. It includes the improved use of fire for sustainable forestry management, including wildfire prevention and prescribed burning. Prescribed burning is used to reduce the risk of large, uncontrollable fires in forest areas, and controlled burning is among the most effective and economic methods of reducing fire danger and stimulating natural reforestation under the forest canopy and after clear felling. | The frequency and severity of large wildfires have increased around the globe in recent decades, which has impacted forest carbon budgets. Fire can cause various greenhouse gas emissions such as CO ₂ , CH ₄ , and N ₂ O, and others such as CO, volatile organic carbon, and smoke aerosols. Fire management can reduce GHG emissions and can reduce haze pollution, which has significant health and economic impacts. Fire management helps to prevent soil erosion and land degradation and is used in rangelands to conserve biodiversity and to enhance forage quality. | Chapter 2; Cross-Chapter Box 3 on fire and climate; Esteves et al. 2012; FAO 2006; Lin et al. 2017; O'Mara 2012; Rulli et al. 2006; Scasta et al. 2016; Seidl et al. 2014; Smith et al. 2014b; Tacconi 2016; Valendik et al. 2011; Westerling et al. 2006; Whitehead et al. 2008; Yong and Peh 2016 |
| Reduced landslides and natural hazards | Landslides are mainly triggered by human activity (e.g. legal and illegal mining, fire, deforestation) in combination with climate. Management of landslides and natural hazards (e.g. floods, storm surges, droughts) is based on vegetation management (e.g. afforestation) and engineering works (e.g. dams, terraces, stabilisation and filling of erosion gullies). | Management of landslides and natural hazards is important for adaptation and is a very important intervention for managing land degradation, since landslides and natural hazards are among the most severe degradation processes. In countries where mountain slopes are planted with food crops, reduced landslides will help deliver benefits for food security. Most deaths caused due to different disasters have occurred in developing countries, in which poverty, poor education and health facilities, and other aspects of development increase exposure, vulnerability and risk. | IPCC AR5 WG2 Chapter 14; Arnáez J et al. 2015; Campbell 2015; ITPS-FAO 2015; Gariano and Guzzetti 2016; Mal et al. 2018 |
| Reduced pollution including acidification | Management of air pollution is connected to climate change by emission sources of air polluting materials and their impacts on climate, human health, and ecosystems, including agriculture. Acid deposition is one of the many consequences of air pollution, harming trees and other vegetation, as well as being a significant driver of land degradation. Practices that reduce acid deposition include prevention of emissions of nitrogen oxides (NO _x) and sulphur dioxide (SO ₂), which also reduce GHG emissions and | There are a few potential adverse side effects of reduction in air pollution to carbon sequestration in terrestrial ecosystems, because some forms of air pollutants can enhance crop productivity by increasing diffuse sunlight, compared to direct sunlight. Reactive N deposition could also enhance CO ₂ uptake in boreal forests and increase soil carbon pools to some extent. Air pollutants have different impacts on climate depending primarily on the composition, with some aerosols (and clouds seeded by them) increasing the reflection of solar radiation to space leading to net cooling, while others (e.g. black carbon and tropospheric ozone) having a net warming effect. Therefore, control of these different pollutants will have both positive and negative impacts on climate mitigation. | Chapter 2; Anderson et al. 2017; Chum et al. 2013; Carter et al. 2015; Coakley; Maaroufi et al. 2015; Markandya et al. 2018; Melamed and Schmale 2016; Mostofa et al 2016.; Nemet et al. 2010; Ramanathan et al. 2001; Seinfeld and |

| | | | |
|--|---|---|--|
| | <p>other Short-Lived Climate Pollutants (SLCPs). Reductions of SLCPs reduce warming in the near term and the overall rate of warming, which can be crucial for plants that are sensitive to even small increases in temperature. Management of harmful air pollutants such as fine particulate matter (PM_{2.5}) and ozone (O₃) also mitigate the impacts of incomplete fossil fuel combustion and GHG emissions. In addition, management of pollutants such as tropospheric O₃ has beneficial impacts on food production, since O₃ decreases crop production. Control of urban and industrial air pollution would also mitigate the harmful effects of pollution and provide adaptation co-benefits <i>via</i> improved human health. Management of pollution contributes to aquatic ecosystem conservation since controlling air pollution, rising atmospheric CO₂ concentrations, acid deposition, and industrial waste will reduce acidification of marine and freshwater ecosystems.</p> | | <p>Pandis; Smith et al. 2015b; UNEP 2017; Wild et al. 2012 UNEP and WMO 2011; Xu & Ramanathan, 2017; Xu et al., 2013</p> |
| <p>Management of invasive species / encroachment</p> | <p>Agriculture and forests can be diverse but often, much of the diversity is non-native. Invasive species in different biomes have been introduced intentionally or unintentionally through export of ornamental plants or animals, and through the promotion of modern agriculture and forestry. Non-native species tend to be more numerous in larger than in smaller human-modified landscapes (e.g. over 50% of species in an urbanised area or extensive agricultural fields can be non-native). Invasive alien species in the United States cause major environmental damage amounting to almost USD120 billion yr⁻¹. There are approximately 50,000 foreign species and the number is increasing. About 42% of the species on the Threatened or Endangered species lists are at risk primarily because of alien-invasive species. Invasive species can be managed</p> | <p>Exotic species are used in forestry where local indigenous forests cannot produce the type, quantity and quality of forest products required. Planted forests of exotic tree species make significant contributions to the economy and provide multiple products and Nature's Contributions to People. In general, exotic species are selected to have higher growth rates than native species and produce more wood per unit of area and time. In 2015, the total area of planted forest with non-native tree species was estimated to around 0.5 Mkm². Introduced species were dominant in South America, Oceania and Eastern and Southern Africa, where industrial forestry is dominant. The use of exotic tree species has played an important role in the production of roundwood, fibre, firewood and other forest products. The challenge is to manage existing and future plantation forests of alien trees to maximise current benefits, while minimising present and future risks and negative impacts, and without compromising future benefits. In many countries or regions, non-native trees planted for production or other purposes often lead to sharp conflicts of interest when they become invasive, and to negative impacts on Nature's Contributions to People and nature conservation.</p> | <p>Brundu and Richardson 2016; Cossalter and Pye-Smith 2003; Dresner et al. 2015; Payn et al. 2015; Pimentel et al. 2005; Vilà et al. 2011</p> |

| | | | |
|--|--|---|---|
| | through manual clearance of invasive species, while in some areas, natural enemies of the invasive species are introduced to control them. | | |
| Restoration and reduced conversion of coastal wetlands | Coastal wetland restoration involves restoring degraded / damaged coastal wetlands including mangroves, salt marshes and seagrass ecosystems. | Coastal wetland restoration and avoided coastal wetland impacts have the capacity to increase carbon sinks and can provide benefits by regulating water flow and preventing downstream flooding. Coastal wetlands provide a natural defence against coastal flooding and storm surges by dissipating wave energy, reducing erosion and by helping to stabilise shore sediments. Since large areas of global coastal wetlands are degraded, restoration could provide benefits land degradation. Since some areas of coastal wetlands are used for food production, restoration could displace food production and damage local food supply (Section 6.3.4), though some forms (e.g. mangrove restoration) can improve local fisheries. | Griscom et al. 2017a; Lotze et al. 2006; Munang et al. 2014; Naylor et al. 2000 |
| Restoration and reduced conversion of peatlands | Peatland restoration involves restoring degraded / damaged peatlands which both increases carbon sinks, but also avoids ongoing CO ₂ emissions from degraded peatlands, so it both prevents future emissions and creates a sink, as well as protecting biodiversity. | Avoided peat impacts and peatland restoration can provide significant mitigation, though restoration can lead to an increase in methane emissions, particularly in nutrient rich fens. There may also be benefits for climate adaptation by regulating water flow and preventing downstream flooding. Considering that large areas of global peatlands are degraded, peatland restoration is a key tool in addressing land degradation. Since large areas of tropical peatlands and some northern peatlands have been drained and cleared for food production, their restoration could displace food production and damage local food supply, potentially leading to adverse impacts on food security locally, though the global impact would be limited due to the relatively small areas affected. | Griscom et al. 2017a; Jauhiainen et al. 2008; Limpens et al. 2008; Munang et al. 2014 |
| Biodiversity conservation | Biodiversity conservation refers to practices aiming at maintaining components of biological diversity. It includes conservation of ecosystems and natural habitats, maintenance and recovery of viable populations of species in their natural surroundings (<i>in-situ</i> conservation) and, in the case of domesticated or cultivated species, in the surroundings where they have developed their distinctive properties outside their natural habitats (<i>ex-situ</i> conservation). Examples of biodiversity conservation measures are establishment of protected areas to achieve specific conservation objectives, preservation of biodiversity hotspots, land management to recover natural habitats, interventions to expand | Biodiversity conservation measures interact with the climate system through many complex processes, which can have either positive or negative impacts. For example, establishment of protected areas can increase carbon storage in vegetation and soil, and tree planting to promote species richness and natural habitats can enhance carbon uptake capacity of ecosystems. Management of wild animals can influence climate <i>via</i> emissions of GHGs (from anaerobic fermentation of plant materials in the rumen), impacts on vegetation (<i>via</i> foraging), changes in fire frequency (as grazers lower grass and vegetation densities as potential fuels), and nutrient cycling and transport (by adding nutrients to soils). Conserving and restoring megafauna in northern regions also prevents thawing of permafrost and reduces woody encroachment, thus avoiding methane emissions and increases in albedo. Defaunation affects carbon storage in tropical forests and savannahs. In the tropics, the loss of mega-faunal frugivores is estimated be responsible for up to 10% reduction in carbon storage of global tropical forests. Frugivore | Bello et al. 2015; Campbell et al. 2008; Crooms et al. 2018; Kapos et al. 2008; Osuri et al. 2016; Schmitz et al. 2018a; Secretariat of the Convention on Biological Diversity 2008 |

| | | | |
|--|---|--|--|
| | or control selective plant or animal species in productive lands or rangelands (e.g., rewilding). | rewilding programmes in the tropics are seen as carbon sequestration options that can be equally effective as tree planting schemes. Biodiversity conservation measures generally favour adaptation, but can interact with food security, land degradation or desertification. Protected areas for biodiversity reduce the land available for food production, and abundancies in some species like large animals can influence land degradation processes by grazing, trampling and compacting soil surfaces, thereby altering surface temperatures and chemical reactions affecting sediment and carbon retention. | |
|--|---|--|--|

1

2

1

Table 6.9 Integrated response options based on land management specifically for CDR

| Integrated response option | Description | Context and caveats | Supporting evidence |
|---------------------------------|---|---|--|
| Enhanced weathering of minerals | The enhanced weathering of minerals that naturally absorb CO ₂ from the atmosphere has been proposed as a CDR technology with a large mitigation potential. The rocks are ground to increase the surface area and the ground minerals are then applied to the land where they absorb atmospheric CO ₂ . | Enhanced mineral weathering can remove atmospheric CO ₂ . Since ground minerals can increase pH, there could be some benefits for efforts to prevent or reverse land degradation where acidification is the driver of degradation. Since increasing soil pH in acidified soils can increase productivity, the same effect could provide some benefit for food security. Minerals used for enhanced weathering need to be mined, and mining has large impacts locally, though the total area mined is likely to be small on the global scale. | Lenton 2010; Schuiling and Krijgsman 2006; Smith et al. 2016a; Taylor et al. 2016a; Beerling et al. 2018 |
| Bioenergy and BECCS | Bioenergy production can mitigate climate change by delivering an energy service, therefore avoiding combustion of fossil energy. It is the most common renewable energy source used today in the world and has a large potential for future deployment (see Cross-Chapter Box 7 on bioenergy in this chapter). BECCS entails the use of bioenergy technologies (e.g. bioelectricity or biofuels) in combination with CO ₂ capture and storage (see also Glossary). BECCS simultaneously provides energy and can reduce atmospheric CO ₂ concentrations (see Chapter 2 and Cross-Chapter Box 7 on bioenergy in this chapter) for a discussion of potentials and atmospheric | Bioenergy and BECCS can compete for land and water with other uses. Increased use of bioenergy and BECCS can result in large expansion of cropland area, significant deforestation, and increased irrigation water use and water scarcity. Large-scale use of bioenergy can result in increased food prices and can lead to an increase in the population at risk of hunger. As a result of these effects, large-scale bioenergy and BECCS can have negative impacts for food security. Interlinkages of bioenergy and BECCS with climate change adaptation, land degradation, desertification, and biodiversity are highly dependent on local factors such as the type of energy crop, management practice, and previous land use. For example, intensive agricultural practices aiming to achieve high crop yields, as is the case for some bioenergy systems, may have significant effects on soil health, including depletion of soil organic matter, resulting in negative impacts on land degradation and desertification. However, with low inputs of fossil fuels and chemicals, limited irrigation, heat/drought tolerant species, using marginal land, biofuel programs can be beneficial to future adaptation of ecosystems. Planting bioenergy crops, like perennial grasses, on degraded land can increase soil carbon and ecosystem quality (including biodiversity), thereby helping to preserve soil quality, reverse land degradation, prevent desertification processes, and reduce food insecurity. These effects depend on the scale of deployment, the feedstock, the prior land use, and which other response options are included (see Section 6.4.4.2). Large-scale production of | Cross-Chapter Box 7 on Bioenergy in this chapter; IPCC SR15; Chapter 2; Chapter 4; Section 6.4; Chapter 7; Baker et al. 2019a; Calvin et al. 2014c; Chaturvedi et al. 2013; Chum et al. 2011; Clarke et al. 2014a; Correa et al. 2017; Creutzig et al. 2015; Dasgupta et al. 2014; Don et al. 2012; Edelenbosch et al. 2017; Edenhofer et al. 2011; FAO 2011; Favero and Mendelsohn 2014; Fujimori et al. 2018a; Fuss et al. 2016, 2018a; Hejazi et al. 2015a; Kemper 2015; Kline et al. 2017; Lal 2014; Lotze-Campen et al. 2013; Mello et al. 2014b; Muratori et al. 2016; Noble et al. 2014; Obersteiner et al. 2016a; Popp et al. 2011c, 2014a, 2017; Riahi et al. 2017a; Robertson et al. 2017b; Sánchez et al. 2017; Searchinger et al. 2018; Sims et al. 2014; Slade et al. 2014; Smith |

| | | | |
|--|---|--|---|
| | <p>effects); thus, BECCS is considered a CDR technology. While several BECCS demonstration projects exist, it has yet to be deployed at scale. Bioenergy and BECCS are widely-used in many future scenarios as a climate change mitigation option in the energy and transport sector, especially those scenarios aimed at a stabilisation of global climate at 2°C or less above pre-industrial levels.</p> | <p>bioenergy can require significant amounts of land, increasing potential pressures for land conversion and land degradation. Low levels of bioenergy deployment require less land, leading to smaller effects on forest cover and food prices; however, these land requirements could still be substantial. In terms of feedstocks, some feedstocks, grown in some regions, may not need irrigation, and thus would not compete for water with food crops. Additionally, the use of residues or microalgae could limit competition for land and biodiversity loss; however, residues could result in land degradation or decreased soil organic carbon. Whether woody bioenergy results in increased competition for land or not is disputed in the literature, with some studies suggesting reduced competition and others suggesting enhanced. One study noted that this effect changes over time, with complementarity between woody bioenergy and forest carbon sequestration in the near-term, but increased competition for land with afforestation/reforestation in the long-term. Additionally, woody bioenergy could also result in land degradation.</p> | <p>et al. 2016a; Torvanger 2018; van Vuuren et al. 2011, 2015b, 2016; Wise et al. 2015; Tian et al. 2018;</p> |
|--|---|--|---|

1

1 **6.2.2 Integrated response options based on value chain management**

2 **6.2.2.1 *Integrated response options based on value chain management through demand***
3 ***management***

4 Integrated response options based on value chain management through demand management are
5 described in Table 6.10, which also notes any context specificities in the effects of the response
6 options and provides the evidence base.

7 **6.2.2.2 *Integrated response options based on value chain management through supply***
8 ***management***

9 Integrated response options based on value chain management through supply management are
10 described in Table 6.11, which also notes any context specificities in the effects of the response
11 options and provides the evidence base.

12 **6.2.3 Integrated response options based on risk management**

13 **6.2.3.1 *Risk management options***

14 Integrated response options based on risk management described in Table 6.12, which also notes any
15 context specificities in the effects of the response options and provides the evidence base.

1

Table 6.10 Integrated response options based on value chain management through demand management

| Integrated response option | Description | Context and caveats | Supporting evidence |
|-----------------------------|---|---|--|
| Dietary change | Sustainable healthy diets represent a range of dietary changes to improve human diets, to make them healthy in terms of the nutrition delivered, and also (economically, environmentally and socially) sustainable. A “contract and converge” model of transition to sustainable healthy diets would involve a reduction in overconsumption (particularly of livestock products) in over-consuming populations, with increased consumption of some food groups in populations where minimum nutritional needs are not met. Such a conversion could result in a decline in undernourishment, as well as reduction in the risk of morbidity and mortality due to over-consumption. | A dietary shift away from meat can reduce greenhouse gas emissions, reduce cropland and pasture requirements, enhance biodiversity protection, and reduce mitigation costs. Additionally, dietary change can both increase potential for other land-based response options and reduce the need for them by freeing land. By decreasing pressure on land, demand reduction through dietary change could also allow for decreased production intensity, which could reduce soil erosion and provide benefits to a range of other environmental indicators such as deforestation and decreased use of fertiliser (N and P), pesticides, water and energy, leading to potential benefits for adaptation, desertification, and land degradation. | Chapter 5; Section 6.4.4.2; Aleksandrowicz et al. 2016a; Bajželj et al. 2014; Bonsch et al. 2016; Erb et al. 2016; Godfray et al. 2010; Haberl et al. 2011; Havlík et al. 2014; Muller et al. 2017a; Smith et al. 2013; Springmann et al. 2018a; Stehfest et al. 2009; Tilman and Clark 2014; Wu et al. 2019 |
| Reduced post-harvest losses | Approximately one-third of the food produced for human consumption is wasted in post-production operations. Most of these losses are due to poor storage management. Post-harvest food losses underlie the food system’s failure to equitably enable accessible and affordable food in all countries. Reduced post-harvest food losses can improve food security in developing countries (while food loss in developed countries mostly occurs at the retail/consumer stage). The key drivers for post-harvest waste in developing countries are structural and infrastructure deficiencies. Thus, reducing food waste at the post-harvest stage requires responses that process, | Differences exist between farm food waste reduction technologies between small-scale agricultural systems and large-scale agricultural systems. A suite of options includes farm level storage facilities, trade or exchange processing technologies including food drying, onsite farm processing for value addition, and improved seed systems. For large scale agri-food systems, options include cold chains for preservation, processing for value addition and linkages to value chains that absorb the harvests almost instantly into the supply chain. In addition to the specific options to reduce food loss and waste, there are more systemic possibilities related to food systems. Improving and expanding the ‘dry chain’ can significantly reduce food losses at the household level. Dry chains are analogous to the | Chapter 5; Anisah et al. 2017; Bajželj et al. 2014; Billen et al. 2018; Bradford et al. 2018; Chaboud and Daviron 2017; Göbel et al. 2015; Gustavsson et al. 2011; Hengsdijk and de Boer 2017; Hodges et al. 2011; Ingram et al. 2016; |

| | | | |
|---|--|---|--|
| | preserve and, where appropriate, redistribute food to where it can be consumed immediately. | cold chain and refers to the ‘initial dehydration of durable commodities to levels preventing fungal growth’ followed by storage in moisture-proof containers. Regional and local food systems are now being promoted to enable production, distribution, access and affordability of food. Reducing post-harvest losses has the potential to reduce emissions and could simultaneously reduce food costs and increase availability. The perishability and safety of fresh foods are highly susceptible to temperature increase. | Kissinger et al. 2018; Kumar and Kalita 2017; Ritzema et al. 2017; Sheahan and Barrett 2017; Wilhelm et al. 2016 |
| Reduced food waste (consumer or retailer) | Since approximately 9-30% of all food is wasted, reducing food waste can reduce pressure on land (see also reducing post-harvest losses). | Reducing food waste could lead to a reduction in cropland area and GHG emissions, resulting in benefits for mitigation. By decreasing pressure on land, food waste reduction could allow for decreased production intensity, which could reduce soil erosion and provide benefits to a range of other environmental indicators such as deforestation and decreases in use of fertiliser (N and P), pesticides, water and energy, leading to potential benefits for adaptation, desertification, and land degradation. | Alexander et al. 2016; Bajželj et al. 2014; Gustavsson et al. 2011; Kummu et al. 2012a; Muller et al. 2017a; Smith et al. 2013; Vermeulen et al. 2012b |
| Material substitution | Material substitution involves the use of wood or agricultural biomass (e.g. straw bales) instead of fossil fuel-based materials (e.g. concrete, iron, steel, aluminium) for building, textiles or other applications. | Material substitution reduces carbon emissions both because the biomass sequesters carbon in materials while re-growth of forests can lead to continued sequestration, and because it reduces the demand for fossil fuels, delivering a benefit for mitigation. However, a potential trade-off exists between conserving carbon stocks and using forests for wood products. If the use of material for substitution was large enough to result in increased forest area, then the adverse side-effects for adaptation and food security would be similar to that of reforestation and afforestation. In addition, some studies indicate that wooden buildings, if properly constructed, could reduce fire risk compared to steel, creating a co-benefit for adaptation. The effects of material substitution on land degradation depend on management practice; some forms of logging can lead to increased land degradation. Long-term forest management with carbon storage in long-lived products also results in atmospheric carbon dioxide (CO ₂) removal. | Chapter 4; Dugan et al. 2018; Eriksson et al. 2012; Gustavsson et al. 2006; Kauppi et al. 2018; Leskinen et al. 2018; McLaren 2012; Oliver and Morecroft 2014; Ramage et al. 2017; Sathre and O’Connor 2010; Smyth et al. 2014; Kurz et al. 2016; Miner 2010; Jordan et al. 2018 |

1

2

1

Table 6.11 Integrated response options based on value chain management through supply management

| Integrated response option | Description | Context and caveats | Supporting evidence |
|-----------------------------|---|---|---|
| Sustainable sourcing | Sustainable sourcing includes approaches to ensure that the production of goods is done in a sustainable way, such as through low-impact agriculture, zero-deforestation supply chains, or sustainably harvested forest products. Currently around 8% of global forest area has been certified in some manner, and 25% of global industrial roundwood comes from certified forests. Sustainable sourcing also aims to enabling producers to increase their percentage of the final value of commodities. Adding value to products requires improved innovation, coordination and efficiency in the food supply chain, as well as labelling to ensure consumer demands. As such, sustainable sourcing is an approach that combines both supply and demand-side management. Promoting sustainable and value-added products can reduce the need for compensatory extensification of agricultural areas and is a specific commitment of some sourcing programs (such as forest certification programs). Table 7.3 (Chapter 7) provides examples of the many sustainable sourcing programs now available globally. | Sustainable sourcing is expanding but accounts for only a small fraction of overall food and material production; many staple food crops do not have strong sustainability standards. Sustainable sourcing provides potential benefits for both climate mitigation and climate adaptation by reducing drivers of unsustainable land management, and by diversifying and increasing flexibility in the food system to climate stressors and shocks. Sustainable sourcing can lower expenditures of food processors and retailers by reducing losses. Adding value to products can extend a producer's marketing season and provide unique opportunities to capture niche markets thereby increasing their adaptive capacity to climate change. Sustainable sourcing can also provide significant benefits for food security, while simultaneously creating economic alternatives for the poor. Sustainable sourcing programmes often also have positive impacts on the overall efficiency of the food supply chain and can create closer and more direct links between producers and consumers. In some cases, processing of value-added products could lead to higher emissions or demand of resources in the food system, potentially leading to small adverse impacts on land degradation and desertification challenges. | Chapter 2; Chapter 3; Chapter 5; Section 6.4; Accorsi et al. 2017; Bajželj et al. 2014; Bustamante et al. 2014a; Clark and Tilman 2017; Garnett 2011; Godfray et al. 2010a; Hertel 2015; Ingram et al. 2016a; James and James 2010a; Muller et al. 2017a; Tilman and Clark 2014a; Springer et al. 2015; Tayleur et al. 2017 |
| Management of supply chains | Management of supply chains include a set of polycentric governance processes focused on improving efficiency and sustainability across the supply chain for each product, to reduce climate risk and profitably reduce emissions. Trade-driven food supply chains are becoming increasingly complex and contributing to emissions. Improved management of supply chains can include both: 1) better food transport and increasing the economic value or reduce risks | Successful implementation of supply chain management practices is dependent on organisational capacity, the agility and flexibility of business strategies, the strengthening of public-private policies and effectiveness of supply-chain governance. Existing practices include a) greening supply chains (e.g. utilising products and services with a reduced impact on the environment and human health), b) adoption of specific sustainability instruments among agri-food companies (e.g. eco- | Chapter 5; Barthel and Isendahl 2013; Haggblade et al. 2017; Lewis and Witham 2012; Michellini et al. 2018; Minot 2014; Mundler and Rumpus 2012; Tadasse et al. |

| | | | |
|--|--|---|--|
| | of commodities through production processes (e.g., packaging, processing, cooling, drying, extracting) and 2) improved policies for stability of food supply, as globalised food systems and commodity markets are vulnerable to food price volatility. The 2007-2008 food price shocks negatively affected food security for millions, most severely in Sub-Saharan Africa. Increasing the stability of food supply chains is a key goal to increase food security, given that climate change threatens to lead to more production shocks in the future. | innovation practices), c) adopting emission accounting tools (e.g. carbon and water foot-printing), and d) implementing “demand forecasting” strategies (e.g. changes in consumer preference for 'green' products). In terms of food supply, measures to improve stability in traded markets can include: 1) financial and trade policies, such as reductions on food taxes and import tariffs; 2) shortening food supply chains (SFSCs); 3) increasing food production; 4) designing alternative distribution networks; 5) increasing food market transparency and reducing speculation in futures markets; 6) increasing storage options; and 7) increasing subsidies and food-based safety nets. | 2016; Wheeler and von Braun 2013; Wilhelm et al. 2016; Wodon and Zaman 2010; The World Bank 2011 |
| Enhanced urban food systems | Urban areas are becoming the principal territories for intervention in improving food access through innovative strategies that aim to reduce hunger and improve livelihoods. Interventions include Urban and Peri-urban Agriculture and Forestry and local food policy and planning initiatives such as Food Policy Councils and city-region-wide regional food strategies. Such systems have demonstrated inter-linkages of the city and its citizens with surrounding rural areas to create sustainable, and more nutritious food supplies for the city, while improving the health status of urban dwellers, reducing pollution levels, adapting to and mitigating climate change, and stimulating economic development. Options include support for urban and peri-urban agriculture, green infrastructure (e.g., green roofs), local markets, enhanced social (food) safety nets and development of alternative food sources and technologies, such as vertical farming. | Urban territorial areas have a potential to reduce GHG emissions through improved food systems to reduce vehicle miles of food transportation, localised carbon capture and food waste reduction. The benefits of Urban food forests that are intentionally planted woody perennial food producing species, are also cited for their carbon sequestration potentials. However, new urban food systems may have diverse unexpected adverse side-effects with climate systems, such as lower efficiencies in food supply and higher costs than modern large-scale agriculture. Diversifying markets, considering value added products in the food supply system may help to improve food security by increasing its economic performance and revenues to local farmers. | Akhtar et al. 2016; Benis and Ferrão 2017; Brinkley et al. 2013, 2016; Chappell et al. 2016; Goldstein et al. 2016; Kowalski and Conway 2018; Lee-Smith 2010; Barthel and Isendahl 2013; Lwasa et al. 2014, 2015; Revi et al. 2014; Specht et al. 2014; Tao et al. 2015; UPAF (date) |
| Improved food processing and retailing | Improved food processing and retailing involves several practices related to a) greening supply chains (e.g., utilising products and services with a reduced impact on the environment and human health), b) adoption of specific sustainability instruments among agri-food companies (e.g., eco-innovation practices), c) adopting emission accounting tools (e.g., carbon and water foot-printing), d) implementing “demand forecasting” strategies (e.g., changes in consumer | Improved food processing and retailing can provide benefits for climate mitigation since GHG-friendly foods can reduce agri-food GHG emissions from transportation, waste and energy use. In cases where climate extremes and natural disasters disrupt supply chain networks, improved food processing and retailing can benefit climate adaptation by buffering the impacts of changing temperature and rainfall patterns on upstream agricultural production. It can provide benefits for food security | Chapter 2; Chapter 5; Avetisyan et al. 2014; Garnett et al. 2013; Godfray et al. 2010; Mohammadi et al. 2014; Porter et al. 2016; Ridoutt et al. 2016; Song et al. 2017 |

| | | | |
|-------------------------------------|---|--|--|
| | preference for 'green' products) and, e) supporting polycentric supply-chain governance processes. | by supporting healthier diets and reducing food loss and waste. Successful implementation is dependent on organisational capacity, the agility and flexibility of business strategies, the strengthening of public-private policies and effectiveness of supply-chain governance. | |
| Improved energy use in food systems | Energy efficiency of agriculture can be improved to reduce the dependency on non-renewable energy sources. This can be realised either by decreased energy inputs, or through increased outputs per unit of input. In some countries, managerial inefficiency (rather than a technology gap) is the main source for energy efficiency loss. Heterogenous patterns of energy efficiency exist at the national scale and promoting energy efficient technologies along with managerial capacity development can reduce the gap and provide large benefits for climate adaptation. Improvements in carbon monitoring and calculation techniques such as the foot-printing of agricultural products can enhance energy efficiency transition management and uptake in agricultural enterprises. | Transformation to low carbon technologies such as renewable energy and energy efficiency can offer opportunities for significant climate change mitigation by providing a substitute to transport fuel (for example) that could benefit marginal agricultural resources, while simultaneously contributing to long term economic growth. In poorer nations, increased energy efficiency in agricultural value-added production, in particular, can provide large mitigation benefits. Under certain scenarios, the efficiency of agricultural systems can stagnate and could exert pressure on grasslands and rangelands, thereby impacting land degradation and desertification. Rebound effects can also occur, with adverse impacts on emissions. | Al-Mansour F and Jejeic V 2017; Baptista et al. 2013; Gunatilake et al. 2014; Begum et al. 2015; Jebli and Youssef 2017; van Vuuren et al. 2017b |

1

2

1

Table 6.12 Integrated response options based on risk management

| Integrated response option | Description | Context and caveats | Supporting evidence |
|----------------------------|--|---|---|
| Management of urban sprawl | Unplanned urbanisation leading to sprawl and extensification of cities along the rural-urban fringe has been identified as a driver of forest and agricultural land loss and a threat to food production around cities. It has been estimated that urban expansion will result in a 1.8–2.4% loss of global croplands by 2030. This rapid urban expansion is especially strong in new emerging towns and cities in Asia and Africa. Policies to prevent such urbanisation have included integrated land use planning, agricultural zoning ordinances and agricultural districts, urban redevelopment, arable land reclamation, and transfer/purchase of development rights or easements. | The prevention of uncontrolled urban sprawl may provide adaptation co-benefits, but adverse side effects for adaptation might arise due to restricted ability of people to move in response to climate change. | Barbero-Sierra et al. 2013a; Bren d'Amour et al. 2016; Cai et al. 2013; Chen 2007; Francis et al. 2012a; Gibson et al. 2015; Lee et al. 2015; Qian et al. 2015; Shen et al. 2017; Tan et al. 2009 |
| Livelihood diversification | When households' livelihoods depend on a small number of sources of income without much diversification, and when those income sources are in fields that are highly climate dependent, like agriculture and fishing, this dependence can put food security and livelihoods at risk. Livelihood diversification (drawing from a portfolio of dissimilar sources of livelihood as a tool to spread risk) has been identified as one option to increase incomes and reduce poverty, increase food security, and promote climate resilience and risk reduction. | Livelihood diversification offers benefits for desertification and land degradation, particularly through non-traditional crops or trees in agroforestry systems which improve soil. Livelihood diversification may increase on-farm biodiversity due to these investments in more ecosystem-mimicking production systems, like agroforestry and polycultures. Diversification into non-agricultural fields, such as wage labour or trading, is increasingly favoured by farmers as a low-cost strategy, particularly to respond to increasing climate risks. | Adger 1999; Ahmed and Stepp 2016a; Antwi-Agyei et al. 2014; Barrett et al. 2001; Berman et al. 2012; Bryceson 1999; DiGiano and Racelis 2012; Ellis 1998, 2008; Ngigi et al. 2017; Rakodi 1999; Thornton and Herrero 2014; Little et al. 2001 |
| Use of local seeds | Using local seeds (also called seed sovereignty) refers to use of non-improved, non-commercial seeds varieties. These can be used and stored by local farmers as low-cost inputs and can often help contribute to the conservation of local varieties and land races, increasing local biodiversity. Many local seeds also require no pesticide or fertiliser use, leading to less land degradation in their use. | Use of local seeds is important in the many parts of the developing world that do not rely on commercial seed inputs. Promotion of local seed saving initiatives can include seed networks, banks and exchanges, and non-commercial open source plant breeding. These locally developed seeds can both help protect local agrobiodiversity and can often be more climate resilient than generic commercial varieties, although | Bowman 2015; Campbell and Veteto 2015; Coomes et al. 2015; Kloppenberg 2010; Luby et al. 2015; van Niekerk and Wynberg 2017; Patnaik et al. 2017; Reisman 2017; |

| | | | |
|--------------------------|---|---|---|
| | | the impacts on food security and overall land degradation are inconclusive. | Vasconcelos et al. 2013; Wattnem 2016 |
| Disaster risk management | Disaster risk management encompasses many approaches to try to reduce the consequences of climate and weather-related disasters and events on socio-economic systems. The Hyogo Plan of Action is a UN framework for nations to build resilience to disasters through effective integration of disaster risk considerations into sustainable development policies. For example, in Vietnam a national strategy on disasters based on Hyogo has introduced the concept of a “four-on-the-spot” approach for disaster risk management of: proactive prevention; timely response; quick and effective recovery; and sustainable development. Other widespread approaches to disaster risk management include using early warning systems that can encompass 1) education systems; 2) hazard and risk maps; 3) hydrological and meteorological monitoring (such as flood forecasting or extreme weather warnings); and 4) communications systems to pass on information to enable action. These approaches have long been considered to reduce the risk of household asset damage during one-off climate events and are increasingly being combined with climate adaptation policies. | Community-based disaster risk management has been pointed to as one of the most successful ways to ensure information reaches people, who need to be participants in risk reduction. Effective disaster risk management approaches must be ‘end-to-end,’ both reaching communities at risk and supporting and empowering vulnerable communities to take appropriate action. The most effective early warning systems are not simply technical systems of information dissemination, but utilise and develop community capacities, create local ownership of the system, and are based on a shared understanding of needs and purpose. Tapping into existing traditional or local knowledge has also been recommended for disaster risk management approaches to reducing vulnerability. | Ajibade and McBean 2014; Alessa et al. 2016; Bouwer et al. 2014; Carreño et al. 2007; Cools et al. 2016; Djalante et al. 2012; Garschagen 2016; Maskrey 2011; Mercer 2010; Sternberg and Batbuyan 2013; Thomalla et al. 2006; Vogel and O’Brien 2006; Schipper and Pelling 2006 |
| Risk sharing instruments | Risk sharing instruments can encompass a variety of approaches. Intra-household risk pooling is a common strategy in rural communities, such as through extended family financial transfers; one study found 65% of poor households in Jamaica report receiving transfers, and such transfers can account for up to 75% of household income or more after crisis events. Community rotating credit associations (ROSCAs) have long been used for general risk pooling and can be a source of financing to cope with climate variability as well. Credit services have been shown to be important for adaptation actions and risk reduction. Insurance of various kinds is also a form of risk pooling. Commercial crop insurance is one of the most widely used risk-hedging | Locally developed risk pooling measures show general positive impacts on household livelihoods. However, more commercial approaches have mixed effects. Commercial crop insurance is highly subsidised in much of the developed world. Index insurance programmes have often failed to attract sufficient buyers or have remained financially unfeasible for commercial insurance sellers. The overall impact of index insurance on food production supply and access has also not been assessed. Traditional crop insurance has generally been seen as positive for food security as it leads to expansion of agricultural production areas and increased food supply. However, insurance may also ‘mask’ truly risky agriculture and prevent farmers from seeking less risky | Akter et al. 2016; Annan and Schlenker 2015; Claassen et al. 2011; Fenton et al. 2017; Giné et al. 2008; Goodwin and Smith 2003; Hammill et al. 2008; Havemenn and Muccione 2011; Jaworski 2016; Meze-Hausken et al. 2009; Morduch and Sharma 2002; Bhattamishra and Barrett 2010; Peterson 2012; |

| | | | |
|--|---|---|---|
| | financial vehicles, and can involve both traditional indemnity-based insurance that reimburses clients for estimated financial losses from shortfalls, or index insurance that pays out the value of an index (such as weather events) rather than actual losses; the former is more common for large farms in the developed world and the latter for smaller non-commercial farms in developing countries. | production strategies. Insurance can also provide perverse incentives for farmers to bring additional lands into crop production, leading to greater risk of degradation. | Sanderson et al. 2013; Skees and Collier 2012; Smith and Glauber 2012 |
|--|---|---|---|

Cross-Chapter Box 7: Bioenergy and Bioenergy with Carbon Dioxide Capture and Storage (BECCS) in mitigation scenarios

Katherine Calvin (The United States of America), Almut Arneth (Germany), Luis Barioni (Brazil), Francesco Cherubini (Norway/Italy), Annette Cowie (Australia), Joanna House (United Kingdom), Francis X. Johnson (Sweden), Alexander Popp (Germany), Joana Portugal Pereira (United Kingdom), Mark Rounsevell (United Kingdom), Raphael Slade (United Kingdom), Pete Smith (United Kingdom)

Bioenergy and BECCS potential

Using biomass to produce heat, electricity and transport fuels (bioenergy) instead of coal, oil, and natural gas can reduce GHG emissions. Combining biomass conversion technologies with systems that capture CO₂ and inject it into geological formations (bioenergy with carbon dioxide capture and storage (BECCS)) can deliver net negative emissions. The net climate effects of bioenergy and BECCS depend on the magnitude of bioenergy supply chain emissions and land/climate interactions, described further below.

Biomass in 2013 contributed ~60 EJ (10%) to global primary energy⁴ (WBA 2016). In 2011, the IPCC *Special Report on Renewable Energy Sources* concluded that biomass supply for energy could reach 100-300 EJ yr⁻¹ by 2050 with the caveat that the technical potential⁵ cannot be determined precisely while societal preferences are unclear; that deployment depends on “factors that are inherently uncertain”; and that biomass use could evolve in a “sustainable” or “unsustainable” way depending on the governance context (IPCC, 2011). The IPCC WGIII AR5 report noted, in addition, that high deployment levels would require extensive use of technologies able to convert lignocellulosic biomass such as forest wood, agricultural residues, and lignocellulosic crops. The SR15 noted that high levels of bioenergy deployment may result in adverse side-effects for food security, ecosystems, biodiversity, water use, and nutrients (de Coninck et al. 2018).

Although estimates of potential are uncertain, there is *high confidence* that the most important factors determining future biomass supply are land availability and land productivity. These factors are in turn determined by competing uses of land and a myriad of environmental and economic considerations (Dornburg et al. 2010; Batidzirai et al. 2012; Erb et al. 2012; Slade 2014, Searle and Malins 2014). Overlaying estimates of technical potential with such considerations invariably results in a smaller estimate. Recent studies that have attempted to do this estimate that 50-244 EJ biomass could be produced on 0.1-13 Mkm² (Fuss et al. 2018a; Schueler et al. 2016; Searle and Malins 2014; IPCC SR15; Wu et al. 2019; Heck et al. 2018; de Coninck et al. 2018). While preferences concerning economic, social and environmental objectives vary geographically and over time, studies commonly estimate “sustainable” potentials by introducing restrictions intended to protect environmental values and avoid negative effects on poor and vulnerable segments in societies.

Estimates of global geological CO₂ storage capacity are large – ranging from 1680 GtCO₂ to 24000 GtCO₂ (McCollum et al. 2014) – however the potential of BECCS may be significantly constrained by socio-political and technical and geographical considerations, including limits to knowledge and experience (Chapter 6, 7).

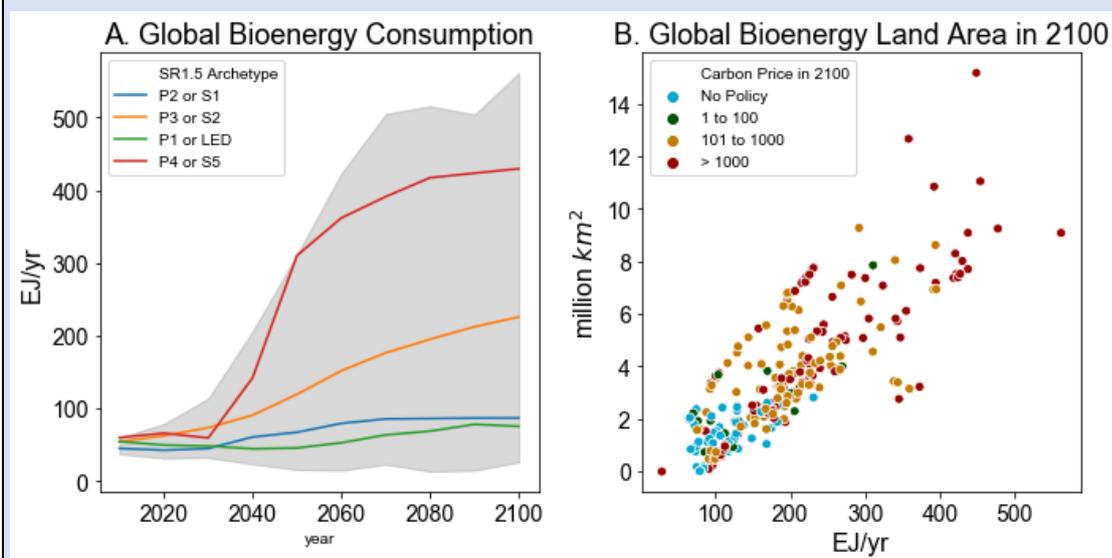
⁴ FOOTNOTE: Of this, more than half was traditional biomass, predominately used for cooking and heating in developing regions, bioelectricity accounted for ~1.7EJ, and transport biofuels for 3.19EJ. (Cross-Chapter Box 12 on Traditional Biomass, Chapter 7)

⁵ FOOTNOTE: The future availability of biomass is usually discussed in terms of a hierarchy of potentials: theoretical>technical>economic. Caution is required, however, as these terms are not always defined consistently and estimates depend on the specific definitions and calculation methodologies.

Bioenergy and BECCS use in mitigation scenarios

Most mitigation scenarios include substantial deployment of bioenergy technologies (Clarke et al. 2014; Fuss et al. 2014; IPCC SR15). Across all scenarios, the amount of bioenergy and BECCS ranges from 0 EJ yr⁻¹ to 561 EJ yr⁻¹ in 2100 (Cross-Chapter Box 7 Figure 1, left panel). Notably, all 1.5°C pathways include bioenergy, requiring as much as 7 Mkm² to be dedicated to the production of energy crops in 2050 (Rogelj et al. 2018a). If BECCS is excluded as a mitigation option, studies indicate that more biomass may be required in order to substitute for a greater proportion of fossil fuels (Muratori et al. 2016; Rose et al. 2014a).

Different Integrated Assessment Models (IAMs) use alternative approaches to land allocation when determining where and how much biomass is used, with some relying on economic approaches and some relying on rule-based approaches (Popp et al. 2014b). Despite these differences a consistent finding across models is that increasing biomass supply to the extent necessary to support deep decarbonisation is likely to involve substantial land use change (Popp et al. 2017) (Cross-Chapter Box 9). In model runs, bioenergy deployment and the consequent demand for biomass and land, is influenced by assumptions around the price of bioenergy, the yield of bioenergy crops, the cost of production (including the costs of fertiliser and irrigation if used), the demand for land for other uses, and the inclusion of policies (e.g., subsidies, taxes, constraints) that may alter land use or bioenergy demand. In general, higher carbon prices result in greater bioenergy deployment (Cross-Chapter Box 7 Figure 1, right panel) and a larger percentage of BECCS. Other factors can also strongly influence bioenergy use, including the cost and availability of fossil fuels (Calvin et al. 2016a), socioeconomics (Popp et al. 2017), and policy (Calvin et al. 2014a; Reilly et al. 2012a).



Cross-Chapter Box 7 Figure 1: Global bioenergy consumption in IAM scenarios. Data is from an update of the IAMC Scenario Explorer developed for the SR15 (Huppmann et al. 2018; Rogelj et al. 2018a). The left panel shows bioenergy deployment over time for the entire scenario database (grey areas) and the four illustrative pathways from SR15 (Rogelj et al. 2018a). The right panel shows global land area for energy crops in 2100 versus total global bioenergy consumption in 2100; colours indicate the carbon price in 2100 (in 2010 USD per tCO₂). Note that this figure includes 409 scenarios, many of which exceed 1.5°C.

Co-benefits, adverse side effect, and risks associated with bioenergy

The production and use of biomass for bioenergy can have co-benefits, adverse side effects, and risks for land degradation, food insecurity, GHG emissions, and other environmental goals. These impacts are context specific and depend on the scale of deployment, initial land use, land type, bioenergy feedstock, initial carbon stocks, climatic region and management regime (Qin et al. 2016; Del Grosso et al. 2014; Alexander et al. 2015; Popp et al. 2017; Davis et al. 2013a; Mello et al. 2014b; Hudiburg

1 et al. 2015; Carvalho et al. 2016; Silva-Olaya et al. 2017; Whitaker et al. 2018; Robledo-Abad et al.
2 2017; Jans et al. 2018).

3 Synergistic outcomes with bioenergy are possible, for example, strategic integration of perennial
4 bioenergy crops with conventional crops can provide multiple production and environmental benefits
5 including management of dryland salinity, enhanced biocontrol and biodiversity, and reduced
6 eutrophication (Davis et al. 2013b; Larsen et al. 2017; Cacho et al. 2018; Odgaard et al. 2019).
7 Additionally, planting perennial bioenergy crops on low carbon soil could enhance soil carbon
8 sequestration (Bárcena et al. 2014; Schröder et al. 2018; Walter et al. 2015; Robertson et al. 2017a;
9 Rowe et al. 2016; Chadwick et al. 2014; Immerzeel et al. 2014; Del Grosso et al. 2014; Mello et al.
10 2014c; Whitaker et al. 2018). However, large-scale expansion of bioenergy may also result in
11 increased competition for land (DeCicco 2013; Humpenöder et al. 2018a; Bonsch et al. 2016; Harris
12 et al. 2015; Richards et al. 2017; Ahlgren et al. 2017; Bárcena et al. 2014), increased greenhouse gas
13 emissions from land use change and land management, loss in biodiversity, and nutrient leakage
14 (Harris et al. 2018; Harper et al. 2018; Popp et al. 2011c; Wiloso et al. 2016; Behrman et al. 2015;
15 Valdez et al. 2017; Hof et al. 2018). If biomass crops are planted on land with a high carbon stock, the
16 carbon loss due to land conversion may take decades to over a century to be compensated by either
17 fossil fuel substitution or CCS (Harper et al. 2018). Competition for land may be experienced locally
18 or regionally and is one of the determinants of food prices, food security (Popp et al. 2014a; Bailey
19 2013; Pahl-Wostl et al. 2018; Rulli et al. 2016; Yamagata et al. 2018; Franz et al. 2017; Kline et al.
20 2017; Schröder et al. 2018) and water availability (Rulli et al. 2016; Bonsch et al. 2015b; Pahl-Wostl
21 et al. 2018; Bailey 2013; Chang et al. 2016; Bárcena et al. 2014).

22 Experience in countries at quite different levels of economic development (Brazil, Malawi and
23 Sweden) has shown that persistent efforts over several decades to combine improved technical
24 standards and management approaches with strong governance and coherent policies, can facilitate
25 long-term investment in more sustainable production and sourcing of liquid biofuels (Johnson and
26 Silveira 2014). For woody biomass, combining effective governance with active forest management
27 over long time periods can enhance substitution-sequestration co-benefits, such as in Sweden where
28 bioenergy has tripled during the last 40 years (currently providing about 25% of total energy supply)
29 while forest carbon stocks have continued to grow (Lundmark et al. 2014). A variety of approaches
30 are available at landscape level and in national and regional policies to better reconcile food security,
31 bioenergy and ecosystem services, although more empirical evidence is needed (Mudombi et al. 2018;
32 Manning et al. 2015; Kline et al. 2017; Maltsoglou et al. 2014; Lamers et al.).

33 Thus, while there is *high confidence* that the technical potential for bioenergy and BECCS is large,
34 there is also *very high confidence* that this potential is reduced when environmental, social and
35 economic constraints are considered. The effects of bioenergy production on land degradation, water
36 scarcity, biodiversity loss, and food insecurity are scale and context specific (*high confidence*).
37 Large areas of monoculture bioenergy crops that displace other land uses can exacerbate these
38 challenges, while integration into sustainably managed agricultural landscapes can ameliorate them
39 (*medium confidence*).

40 **Inventory reporting for BECCS and bioenergy**

41 One of the complications in assessing the total GHG flux associated with bioenergy under
42 UNFCCC reporting protocols is that fluxes from different aspects of bioenergy life cycle are reported
43 in different sectors and are not linked. In the energy sector, bioenergy is treated as carbon neutral at
44 the point of biomass combustion because all change in land carbon stocks due to biomass harvest or
45 land use change related to bioenergy are reported under AFOLU sector. Use of fertilisers is captured
46 in the Agriculture sector, while fluxes related to transport/conversion and removals due to CCS are
47 reported in the energy sector. IAMs follow a similar reporting convention. Thus, the whole life cycle

GHG effects of bioenergy systems are not readily observed in national GHG inventories or modelled emissions estimates (see also IPCC 2006; SR15 Chapter 2 Technical Annex; Chapter 2).

Bioenergy in this report

Bioenergy and BECCS are discussed throughout this special report. Chapter 1 provides an introduction to bioenergy and BECCS and its links to land and climate. Chapter 2 discusses mitigation potential, land requirements and biophysical climate implications. Chapter 4 includes a discussion of the threats and opportunities with respect to land degradation. Chapter 5 discusses linkages between bioenergy and BECCS and food security. Chapter 6 synthesises the co-benefits and adverse side-effects for mitigation, adaptation, desertification, land degradation, and food security, as well as barriers to implementation (e.g., cost, technological readiness, etc.). Chapter 7 includes a discussion of risk, policy, governance, and decision-making with respect to bioenergy and BECCS.

6.3 Potentials for addressing the land challenges

In this section, we assess how each of the integrated response options described in Section 6.2 address the land challenges of climate change mitigation (6.3.1), climate change adaptation (6.3.2), desertification (6.3.3), land degradation (6.3.4), and food security (6.3.5). The quantified potentials across all of mitigation, adaptation, desertification, land degradation and food security are summarised and categorised for comparison in section 6.3.6.

6.3.1 Potential of the integrated response options for delivering mitigation

In this section, the impacts of integrated response options on climate change mitigation are assessed.

6.3.1.1 Integrated response options based on land management

In this section, the impacts on climate change mitigation of integrated response options based on land management are assessed. Some of the caveats of these potential mitigation studies are discussed in Chapter 2 and section 6.2.1.

6.3.1.1.1 Integrated response options based on land management in agriculture

Increasing the productivity of land used for food production can deliver significant mitigation by avoiding emissions that would occur if increased food demand were met through expansion of the agricultural land area (Burney et al., 2010). If pursued through increased agrochemical inputs, numerous adverse impacts on greenhouse gas emissions (and other environmental sustainability) can occur (Table 6.5), but if pursued through sustainable intensification, increased food productivity could provide high levels of mitigation. For example, yield improvement has been estimated to have contributed to emissions savings of >13 GtCO₂ yr⁻¹ since 1961 (Burney et al., 2010; Table 6.13). This can also reduce the greenhouse gas intensity of products (Bennetzen et al., 2016) which means a smaller environmental footprint of production, since demand can be met using less land and/or with fewer animals.

Improved cropland management could provide moderate levels of mitigation (1.4-2.3 GtCO₂e yr⁻¹; Smith et al. 2008, 2014c; Pradhan et al., 2013; Table 6.13). The lower estimate of potential is from Pradhan et al. (2013) for decreasing emissions intensity, and the upper end of technical potential is estimated by adding technical potentials for cropland management (about 1.4 GtCO₂e yr⁻¹), rice management (about 0.2 GtCO₂e yr⁻¹) and restoration of degraded land (about 0.7 GtCO₂e yr⁻¹) from Smith et al. (2008) and Smith et al. (2014c). Note that much of this potential arises from soil carbon sequestration so there is an overlap with that response option (see 6.3.1.1.3).

Grazing lands can store large stocks of carbon in soil and root biomass compartments (Conant and Paustian 2002; O'Mara 2012; Zhou et al. 2017). The global mitigation potential is moderate (1.4–1.8 GtCO₂ yr⁻¹), with the lower value in the range for technical potential taken from Smith et al. (2008)

1 which includes only grassland management measures, and the upper value in the range from Herrero
2 et al. (2016), which includes also indirect effects and some components of livestock management, and
3 soil carbon sequestration, so there is overlap with these response options (see below and 6.3.1.1.3).
4 Conant et al. (2005) caution that increases in soil carbon stocks could be offset by increases in N₂O
5 fluxes.

6 The mitigation potential of improved livestock management is also moderate (0.2–1.8 GtCO₂e yr⁻¹;
7 Smith et al. (2008) including only direct livestock measures; Herrero et al. (2016) include also indirect
8 effects, and some components of grazing land management and soil carbon sequestration) to high
9 (6.13 Gt CO₂e yr⁻¹; Pradhan et al., 2013; Table 6.13). There is an overlap with other response options
10 (see above and 6.3.1.1.3).

11 Zomer et al. (2017) reported that the trees agroforestry landscapes have increased carbon stock by
12 7.33 GtCO₂ between 2000–2010, which is equivalent to 0.7 GtCO₂ yr⁻¹. Estimates of global potential
13 range from 0.1 GtCO₂ yr⁻¹ to 5.7 GtCO₂ yr⁻¹ (from an optimum implantation scenario of Hawken,
14 2014), based on an assessment of all values in Griscom et al. (2017a), Hawken (2014), Zomer et al
15 2016., and Dickie et al. (2014) (Table 6.13).

16 Agricultural diversification mainly aims at increasing climate resilience, but it may have a small (but
17 globally unquantified) mitigation potential as a function of type of crop, fertiliser management, tillage
18 system, and soil type (Campbell et al. 2014; Cohn et al. 2017).

19 Reducing conversion of grassland to cropland could provide significant climate mitigation by
20 retaining soil carbon stocks that might otherwise be lost. When grasslands are converted to croplands,
21 they lose about 36% of their soil organic carbon stocks after 20 years (Poeplau et al. 2011). Assuming
22 an average starting soil organic carbon stock of grasslands of 115 t C ha⁻¹ (Poeplau et al. 2011), this is
23 equivalent to a loss of 41.5 t C ha⁻¹ on conversion to cropland. Mean annual global cropland
24 conversion rates (1961–2003) have been around 47000 km² yr⁻¹ (Krause et al. 2017), or 940000 km²
25 over a 20 year period. The equivalent loss of soil organic carbon over 20 years would therefore be 14
26 Gt CO₂e = 0.7 Gt CO₂ yr⁻¹. Griscom et al. (2017a) estimate a cost-effective mitigation potential of
27 0.03 Gt CO₂ yr⁻¹ (Table 6.13).

28 Integrated water management provides moderate benefits for climate mitigation due to interactions
29 with other land management strategies. For example, promoting soil carbon conservation (e.g.
30 reduced tillage) can improve the water retention capacity of soils. Jat et al. (2015) found that
31 improved tillage practices and residue incorporation increased water-use efficiency by 30%, rice–
32 wheat yields by 5–37%, income by 28–40% and reduced GHG emission by 16–25%. While irrigated
33 agriculture accounts for only 20% of the total cultivated land, the energy consumption from
34 groundwater irrigation is significant. However, current estimates of mitigation potential are limited to
35 reductions in greenhouse gas emissions mainly in cropland and rice cultivation (Chapter 2; Table
36 6.13; Smith et al. 2008, 2014c). Li et al. (2006) estimated a 0.52–0.72 GtCO₂ yr⁻¹ reduction using the
37 alternate wetting and drying technique. Current estimates of N₂O release from terrestrial soils and
38 wetlands accounts for 10–15% of anthropogenically fixed nitrogen on the Earth System (Wang et al.
39 2017).

40 Table 6.13 summarises the mitigation potentials for agricultural response options, with confidence
41 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
42 exhaustive) references upon which the evidence is based.

43 **Table 6.13 Mitigation effects of response options based on land management in agriculture**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------|--|----------------|-------------------------------|
| Increased food productivity | >13 GtCO ₂ e yr ⁻¹ | Low confidence | Chapter 5; Burney et al. 2010 |

| | | | |
|---|--|-------------------|---|
| Improved cropland management ^a | 1.4-2.3 GtCO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Chapter 5; Smith et al. 2008, 2014c; Pradhan et al., 2013; |
| Improved grazing land management ^a | 1.4-1.8 GtCO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Chapter 5; Conant et al. 2017; Herrero et al. 2016; Smith et al. 2008, 2014c |
| Improved livestock management ^a | 0.2-2.4 GtCO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Chapter 5; Smith et al. 2008, 2014c; Herrero et al. 2016 |
| Agroforestry | 0.1-5.7 Gt CO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Griscom et al. 2017a; Zomer et al. 2016; Dickie et al. 2014; Hawken 2014; |
| Agricultural diversification | > 0 | Low confidence | Campbell et al. 2014; Cohn et al. 2017 |
| Reduced grassland conversion to cropland | 0.03-0.7 Gt CO ₂ e yr ⁻¹ | Low confidence | Note high value not shown in Chapter 2; Calculated from values in Krause et al. 2017 and Poeplau et al. 2011; Griscom et al. 2017 |
| Integrated water management | 0.1-0.72 Gt CO ₂ yr ⁻¹ | Low confidence | IPCC 2014; Smith et al. 2008, 2014b; Howell et al. 2015; Li et al. 2006; Rahman and Bulbul 2015 |

1 ^a Note that Chapter 2 reports mitigation potential for subcategories within this response option and not the combined total
2 reported here.

3 **6.3.1.1.2 Integrated response options based on land management in forests**

4 Improved forest management could potentially contribute to moderate mitigation benefits globally, up
5 to about 2 Gt CO₂e yr⁻¹ (Chapter 2, Table 6.14). For managed forests, the most effective forest carbon
6 mitigation strategy is the one that, through increasing biomass productivity, optimises the carbon
7 stocks (in forests and in long-lived products) as well as the wood substitution effects for a given time
8 frame (Smyth et al. 2014; Grassi et al. 2018; Nabuurs et al. 2007; Lewis et al. 2019) (Kurz et al. 2016;
9 Erb et al. 2018). Estimates of the mitigation potential vary also depending on the counterfactual, such
10 as business-as-usual management (e.g. Grassi et al. 2018) or other scenarios. Climate change will
11 affect the mitigation potential of forest management due to an increase in extreme events like fires,
12 insects and pathogens (Seidl et al. 2017). More detailed estimates are available at regional or biome
13 level. For instance, according to Nabuurs et al. (2017), the implementation of Climate-Smart Forestry
14 (a combination of improved forest management, expansion of forest areas, energy substitution,
15 establishment of forest reserves, etc.) in the European Union has the potential to contribute to an
16 additional 0.4 Gt CO₂ yr⁻¹ mitigation by 2050. Sustainable forest management is often associated with
17 a number of co-benefits for adaptation, ecosystem services, biodiversity conservation, microclimatic
18 regulation, soil erosion protection, coastal area protection and water and flood regulation (Locatelli
19 2011). Forest management mitigation measures are more likely to be long-lasting if integrated into
20 adaptation measures for communities and ecosystems, for example, through landscape management
21 (Locatelli et al. 2011). Adoption of reduced-impact logging and wood processing technologies along
22 with financial incentives can reduce forest fires, forest degradation, maintain timber production, and
23 retain carbon stocks (Sasaki et al. 2016). Forest certification may support sustainable forest
24 management, helping to prevent forest degradation and over-logging (Rametsteiner and Simula 2003).
25 Community forest management has proven a viable model for sustainable forestry, including for
26 carbon sequestration (Chhatre and Agrawal 2009, Chapter 7, section 7.6.4).

27 Reducing deforestation and forest degradation rates represents one of the most effective and robust
28 options for climate change mitigation, with large mitigation benefits globally (Chapter 2, Chapter 4,
29 Table 6.14). Because of the combined climate impacts of GHGs and biophysical effects, reducing
30 deforestation in the tropics has a major climate mitigation effect, with benefits at local levels too

1 (Chapter 2, Alkama and Cescatti 2016). Reduced deforestation and forest degradation typically lead to
2 large co-benefits for other ecosystem services (Table 6.14).

3 A large range of estimates exist in the scientific literature for the mitigation potential of reforestation
4 and forest restoration, and they sometimes overlap with estimates for afforestation. At global level the
5 overall potential for these options is large, reaching about 10 GtCO₂ yr⁻¹ (Chapter 2, Table 6.14). The
6 greatest potential for these options is in tropical and subtropical climate (Houghton and Nassikas
7 2018). Furthermore, climate change mitigation benefits of afforestation, reforestation and forest
8 restoration are reduced at high latitudes owing to the surface albedo feedback (see Chapter 2).

9 Table 6.14 summarises the mitigation potentials for forest response options, with confidence estimates
10 based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not exhaustive)
11 references upon which the evidence is based.

12 **Table 6.14 Mitigation effects of response options based on land management in forests**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|--|-------------------|--|
| Improved forest management | 0.4-2.1 Gt CO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Griscom 2017; Sasaki et al. 2016 |
| Reduced deforestation and degradation | 0.4-5.8 Gt CO ₂ e yr ⁻¹ | High confidence | Chapter 2; Houghton & Nassikas 2018; Griscom 2017; Baccini 2017; Hawken 2017; Houghton et al 2015; Smith et al. 2014a |
| Reforestation and forest restoration | 1.5-10.1 Gt CO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Dooley and Kartha 2018a; Hawken 2017; Houghton & Nassikas 2018; Griscom 2017. Estimates partially overlapping with Afforestation. |
| Afforestation | 0.5-8.9 Gt CO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Fuss et al. 2018; Hawken 2017; Kreidenweis et al. 2016; Lenton 2010. Estimates partially overlapping with Reforestation. |

13

14 **6.3.1.1.3 Integrated response options based on land management of soils**

15 The global mitigation potential for increasing soil organic matter stocks in mineral soils is estimated
16 to be in the range of 1.3–5.1 GtCO₂e yr⁻¹, though the full literature range is wider (Smith et al. 2008;
17 Smith 2016; Fuss et al 2018.; Sanderman et al. 2017; Sommer & Bossio 2014; Lal 2004; Lal et al.
18 2010; Table 6.15).

19 The management and control of erosion may prevent losses of organic carbon in water- or wind-
20 transported sediments, but since the final fate of eroded material is still debated, ranging from a
21 source of 1.36–3.67 GtCO₂ yr⁻¹ (Jacinthe and Lal 2001; Lal et al., 2004) to a sink of 0.44–3.67 GtCO₂
22 yr⁻¹ (Stallard 1998; Smith et al. 2001, 2005; Van Oost et al. 2007; Table 6.15), the overall impact of
23 erosion control on mitigation is context-specific and at the uncertain at the global level (Hoffmann et
24 al., 2013).

1 Salt-affected soils are highly constrained environments that require permanent prevention of
2 salinisation. Their mitigation potential is likely to be small (Wong et al. 2010; UNCTAD 2011; Dagar
3 et al. 2016b).

4 Soil compaction prevention could reduce N₂O emissions by minimising anoxic conditions favourable
5 for denitrification (Mbow et al. 2010), but its carbon sequestration potential depends on crop
6 management and the global mitigation potential, though globally unquantified, is likely to be small
7 (Chamen et al. 2015; Epron et al. 2016; Tullberg et al. 2018; Table 6.15).

8 For biochar, a global analysis of technical potential, in which biomass supply constraints were applied
9 to protect against food insecurity, loss of habitat and land degradation, estimated technical potential
10 abatement of 3.7–6.6 GtCO₂e yr⁻¹ (including 2.6–4.6 GtCO₂e yr⁻¹ carbon stabilisation). Considering
11 all published estimates by Woolf et al. (2010), Smith (2016), Fuss et al. (2018b), Griscom et al.
12 (2017), Hawken (2017), Paustian et al. (2016), Powell & Lenton (2012), Dickie et al. (2014), Lenton
13 (2010), Lenton (2014), Roberts et al. (2010), Pratt & Moran (2010) and IPCC (2018), the low value
14 for the range of potentials of 0.03 GtCO₂e yr⁻¹ is for the “plausible” scenario of Hawken, 2017; Table
15 6.15). Fuss et al. (2018) propose a range of 0.5–2 GtCO₂e yr⁻¹ as the sustainable potential for negative
16 emissions through biochar, similar to the range proposed by Smith (2016) and IPCC (2018).

17 Table 6.15 summarises the mitigation potentials for soil-based response options, with confidence
18 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
19 exhaustive) references upon which the evidence is based.

20

Table 6.15 Mitigation effects of response options based on land management of soils

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|--|-------------------|---|
| Increased soil organic carbon content | 0.4-8.6 GtCO ₂ e yr ⁻¹ | High confidence | Chapter 2; McLaren 2012; Poeplau and Don 2015; Conant et al. 2017; Dickie et al. 2014; Frank et al. 2017; Fuss et al. 2018b; Griscom et al. 2017; Herrero et al. 2016; Paustian et al. 2016; Powlson et al. 2014; Sanderman et al. 2017; Smith 2016b; Zomer et al. 2016; Hawken 2017; Henderson et al. 2015; Lal 2004; Lal et al. 2010; Sommer & Bossio 2014; |
| Reduced soil erosion | Source of 1.36-3.67 to sink of 0.44-3.67 Gt CO ₂ e yr ⁻¹ | Low confidence | Chapter 2; Jacinthe and Lal 2001; Smith et al. 2001, 2005; Stallard 1998; Van Oost et al. 2007; Lal et al., 2004; Stallard, 1998 |
| Reduced soil salinisation | >0 | Low confidence | Dagar et al. 2016b; UNCTAD 2011; Wong et al. 2010 |
| Reduced soil compaction | >0 | Low confidence | Chamen et al. 2015b; Epron et al. 2016; Tullberg et al. 2018b |
| Biochar addition to soil | 0.03-6.6 GtCO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; IPCC 2018; Fuss et al. 2018b; Griscom et al. 2017a; Lenton 2010; Paustian et al. 2016; Smith 2016; Woolf et al. 2010; Dickie et al. 2014; Hawken 2017; Lenton 2014; Powell & Lenton 2012; Pratt & Moran 2010; Roberts et al 2010; |

21

22 **6.3.1.1.4 Integrated response options based on land management in all/other ecosystems**

23 For fire management, total emissions from fires have been in the order of 8.1 GtCO₂e yr⁻¹ for the
24 period 1997-2016 (see Chapter 2, Cross-Chapter Box 3) and there are important synergies between air
25 pollution and climate change control policies. Reduction in fire CO₂ emissions due to fire suppression
26 and landscape fragmentation associated with increases in population density is calculated to enhance

1 land carbon uptake by 0.48 Gt CO₂e yr⁻¹ for the 1960–2009 period (Arora and Melton 2018; Table
2 6.16).

3 Management of landslides and natural hazards is a key climate adaptation option but due to limited
4 global areas vulnerable to landslides and natural hazards, its mitigation potential is likely to be modest
5 (Noble et al. 2015).

6 In terms of management of pollution, including acidification, UNEP and WMO (2011) and Shindell et
7 al. (2012) identified measures targeting reduction in SLCP emissions that reduce projected global
8 mean warming about 0.5°C by 2050. Bala et al. (2013) reported that a recent coupled modelling study
9 showed N deposition and elevated CO₂ could have a synergistic effect, which could explain 47% of
10 terrestrial carbon uptake in the 1990s. Estimates of global terrestrial carbon uptake due to current N
11 deposition ranges between 0.55 and 1.28 GtCO₂ yr⁻¹ (de Vries et al. 2006; de Vries et al. 2009; Bala et
12 al. 2013; Zaehle and Dalmonech 2011; Table 6.16).

13 There are no global data on the impacts of management of invasive species / encroachment on
14 mitigation.

15 Coastal wetland restoration could provide high levels of climate mitigation, with avoided coastal
16 wetland impacts and coastal wetland restoration estimated to deliver 0.3-3.1 GtCO₂e yr⁻¹ in total when
17 considering all global estimates from Griscom et al. (2017a), Hawken (2017), Pendleton et al. (2012),
18 Howard et al. (2017) and Donato et al. (2010) (Table 6.16).

19 Peatland restoration could provide moderate levels of climate mitigation, with avoided peat impacts
20 and peat restoration estimated to deliver 0.6-2 GtCO₂e yr⁻¹ from all global estimates published in
21 Griscom et al. (2017a), Hawken (2017), Hooijer et al. (2010), Couwenberg et al. (2010) and Joosten
22 and Couwenberg (2008), though there could be an increase in methane emissions after restoration
23 (Jauhiainen et al. 2008; Table 6.16).

24 Mitigation potential from biodiversity conservation varies depending on the type of intervention and
25 specific context. Protected areas are estimated to store over 300 Gt carbon, roughly corresponding to
26 15% of terrestrial carbon stocks (Kapos et al. 2008; Campbell et al. 2008). At global level, the
27 potential mitigation resulting from protection of these areas for the period 2005-2095 is on average
28 about 0.9 GtCO₂-eq. yr⁻¹ relative to a reference scenario (Calvin et al. 2014a). The potential effects on
29 the carbon cycle of management of wild animal species are case context dependent. For example,
30 moose browsing in boreal forests can decrease the carbon uptake of ecosystems by up to 75%
31 (Schmitz et al. 2018b), and reducing moose density through active population management in Canada
32 is estimated to be a carbon sink equivalent to about 0.37 Gt CO₂e yr⁻¹ (Schmitz et al. 2014).

33 Table 6.16 summarises the mitigation potentials for land management response options in all/other
34 ecosystems, with confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6,
35 and indicative (not exhaustive) references upon which the evidence is based.

36 **Table 6.16 Mitigation effects of response options based on land management in all/other ecosystems**

| Integrated response option | Potential | Confidence | Citation |
|---|---|-----------------------------|---|
| Fire management | 0.48-8.1 GtCO ₂ e yr ⁻¹ | Medium confidence | Chapter 2, Cross-Chapter Box 3 on Fire (Chapter 2); Arora and Melton 2018; Tacconi 2016 |
| Reduced landslides and natural hazards | >0 | Low confidence | |
| Reduced pollution including acidification | 1) Reduce projected warming ~0.5°C by 2050; 2) Reduce terrestrial C uptake 0.55-1.28 GtCO ₂ e yr ⁻¹ | 1) and 2) Medium confidence | 1) Shindell et al., 2012; UNEP and WMO, 2011; 2) Bala et al. 2013 |

| | | | |
|--|--|-------------------|---|
| Management of invasive species / encroachment | No global estimates | No evidence | |
| Restoration and reduced conversion of coastal wetlands | 0.3-3.1 GtCO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Griscom et al. 2017a; Donato et al. 2010; Hawken 2017; Howard et al. 2017; Pendleton et al. 2012; |
| Restoration and reduced conversion of peatlands | 0.6-2 GtCO ₂ e yr ⁻¹ | Medium confidence | Chapter 2; Couwenberg et al. 2010; Griscom et al. 2017a; Hooijer et al. 2010; Joosten and Couwenberg 2008; Hawken 2017; |
| Biodiversity conservation | ~0.9 GtCO ₂ -e yr ⁻¹ | Low confidence | Chapter 2; Calvin et al. 2014c; Schmitz et al. 2014 |

1

2 **6.3.1.1.5 Integrated response options based on land management specifically for CDR**

3 Enhanced mineral weathering provides substantial climate mitigation, with a global mitigation
4 potential in the region of about 0.5–4 GtCO₂e yr⁻¹ (Beerling et al 2018.; Lenton 2010; Smith et al.
5 2016c; Taylor et al. 2016; Table 6.17).

6 The mitigation potential for bioenergy and BECCS derived from bottom-up models is large (IPCC
7 SR15; Chapter 2; Cross-Chapter Box 7 on Bioenergy in this chapter), with technical potential
8 estimated at 100-300 EJ yr⁻¹ (IPCC 2011; Cross-Chapter Box 7 in this chapter) or up to ~11 GtCO₂ yr⁻¹
9 (Chapter 2). These estimates, however, exclude N₂O associated with fertiliser application and land-
10 use change emissions. Those effects are included in the modelled scenarios using bioenergy and
11 BECCS, with the sign and magnitude depending on where the bioenergy is grown (Wise et al. 2015),
12 at what scale, and whether N fertiliser is used.

13 Table 6.17 summarises the mitigation potentials for land management options specifically for CDR,
14 with confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and
15 indicative (not exhaustive) references upon which the evidence is based.

16 **Table 6.17 Mitigation effects of response options based on land management specifically for CDR**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------|---|-------------------|---|
| Enhanced weathering of minerals | 0.5-4 GtCO ₂ yr ⁻¹ | Medium confidence | Chapter 2; Beerling et al.; Lenton 2010; Smith et al. 2016c; Taylor et al. 2016b |
| Bioenergy and BECCS | 0.4-11.3 GtCO ₂ yr ⁻¹ | Medium confidence | Chapter 2; IPCC SR15; Fuss et al. 2018b; Lenton 2014; McLaren 2012; Lenton 2010; Powell and Lenton 2012 |

17

18 **6.3.1.2 Integrated response options based on value chain management**

19 In this section, the impacts on climate change mitigation of integrated response options based on
20 value chain management are assessed.

21 **6.3.1.2.1 Integrated response options based on value chain management through demand 22 management**

23 Dietary change and waste reduction can provide large benefits for mitigation, with potentials of 0.7-8
24 GtCO₂ yr⁻¹ for both (Bajželj et al. 2014; Stehfest et al. 2009; Tilman and Clark 2014b;
25 Aleksandrowicz et al. 2016; Herrero et al. 2016; Springmann et al. 2016; Smith et al. 2013; Dickie et
26 al. 2014; Popp et al. 2010; Hawken 2017; Hedenus (2014)). Estimates for food waste reduction

1 (Hawken 2017; Hic et al. 2016; Dickie et al. 2014) (Bajželj et al. 2014) include both consumer /
2 retailed waste and post-harvest losses (Table 6.18).

3 Some studies indicate that material substitution has the potential for significant mitigation, with one
4 study estimating a 14–31% reduction in global CO₂ emissions (Oliver et al. 2014); other studies
5 suggest more modest potential (Gustavsson et al. 2006; Table 6.18).

6 Table 6.18 summarises the mitigation potentials for demand management options, with confidence
7 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
8 exhaustive) references upon which the evidence is based.

9 **Table 6.18 Mitigation effects of response options based on demand management**

| Integrated response option | Potential | Confidence | Citation |
|---|---|-------------------|--|
| Dietary change | 0.7 to 8 GtCO ₂ yr ⁻¹ | High confidence | Chapter 2; Chapter 5; Bajželj et al. 2014; Herrero et al. 2016; Smith et al. 2013; Springmann et al. 2016, 2018b; Stehfest et al. 2009; Tilman and Clark 2014b; Dickie et al. 2014; Hawken 2017; Hedenus 2014; Popp et al. 2010; |
| Reduced post-harvest losses | 4.5 GtCO ₂ yr ⁻¹ | High confidence | Chapter 5; Bajželj et al. 2014 |
| Reduced food waste (consumer or retailer) | 0.8 to 4.5 GtCO ₂ yr ⁻¹ | High confidence | Chapter 5; Bajželj et al. 2014; Dickie et al. 2014; Hawken 2017; Hic et al. 2016 |
| Material substitution | 0.25 to 1 GtCO ₂ yr ⁻¹ | Medium confidence | Chapter 2; Dugan et al. 2018; Gustavsson et al. 2006; Leskinen et al. 2018; McLaren 2012; Sathre and O'Connor 2010; Miner 2010; Kauppi 2001; Smyth et al. 2016 |

10

11 **6.3.1.2.2 Integrated response options based on value chain management through supply** 12 **management**

13 While sustainable sourcing presumably delivers a mitigation benefit, there are no global estimates of
14 potential. Palm oil production alone is estimated to contribute 0.038 to 0.045 GtC yr⁻¹, and Indonesian
15 palm oil expansion contributed up to 9% of tropical land use change carbon emissions in the 2000s
16 (Carlson and Curran 2013), however, the mitigation benefit of sustainable sourcing of palm oil has not
17 been quantified. There are no estimates of the mitigation potential for urban food systems.

18 Efficient use of energy and resources in food transport and distribution contribute to a reduction in
19 GHG emissions, estimated to be 1% of global CO₂ emissions (James and James 2010; Vermeulen et
20 al. 2012). Given that global CO₂ emissions in 2017 were 37 GtCO₂, this equates to 0.37 GtCO₂ yr⁻¹
21 (covering food transport and distribution, improved efficiency of food processing and retailing, and
22 improved energy efficiency; Table 6.19).

23 Table 6.19 summarises the mitigation potentials for supply management options, with confidence
24 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
25 exhaustive) references upon which the evidence is based.

26 **Table 6.19 Mitigation effects of response options based on supply management**

| Integrated response option | Potential | Confidence | Citation |
|--|---|----------------|-----------------|
| Sustainable sourcing | No global estimates | No evidence | |
| Management of supply chains | No global estimates | No evidence | |
| Enhanced urban food systems | No global estimates | No evidence | |
| Improved food processing and retailing | See improved energy efficiency | | |
| Improved energy use in food | 0.37 GtCO ₂ yr ⁻¹ | Low confidence | James and James |

| | | | |
|---------|--|--|-------------------------------|
| systems | | | 2010b; Vermeulen et al. 2012b |
|---------|--|--|-------------------------------|

1

2 **6.3.1.3 Integrated response options based on risk management**

3 In this section, the impacts on climate change mitigation of integrated response options based on risk
4 management are assessed. In general, because these options are focused on adaptation and other
5 benefits, the mitigation benefits are modest, and mostly unquantified.

6 Extensive and less dense urban development tends to have higher energy usage, particularly from
7 transport (Liu et al. 2015), such that a 10% reduction of very low density urban fabrics is correlated
8 with 9% fewer emissions per capita in Europe (Baur et al. 2015). However, the exact contribution to
9 mitigation from the prevention of land conversion in particular has not been well quantified
10 (Thornbush et al. 2013). Suggestions from select studies in the US are that biomass decreases by half
11 in cases of conversion from forest to urban land uses (Briber et al. 2015), and a study in Bangkok
12 found a decline by half in carbon sinks in the urban area in the past 30 years (Ali et al. 2018).

13 There is no literature specifically on linkages between livelihood diversification and climate
14 mitigation benefits, although some forms of diversification that include agroforestry would likely
15 result in increased carbon sinks (Altieri et al. 2015; Descheemaeker et al. 2016). There is no literature
16 exploring linkages between local seeds and GHG emission reductions, although use of local seeds
17 likely reduces emissions associated with transport for commercial seeds, though the impact has not
18 been quantified.

19 While disaster risk management can presumably have mitigation co-benefits, as it can help reduce
20 food loss on-farm (e.g. crops destroyed before harvest or avoided animal deaths during droughts and
21 floods meaning reduced production losses and wasted emissions), there is no quantified global
22 estimate for this potential.

23 Risk sharing instruments could have some mitigation co-benefits if they buffer household losses and
24 reduce the need to expand agricultural lands after experiencing risks. However, the overall impacts of
25 these are unknown. Further, commercial insurance may induce producers to bring additional land into
26 crop production, particularly marginal or land with other risks that may be more environmentally
27 sensitive (Claassen et al. 2011). Policies to deny crop insurance to farmers who have converted
28 grasslands in the US resulted in a 9% drop in conversion, which likely has positive mitigation impacts
29 (Claassen et al. 2011). Estimates of emissions from cropland conversion in the US in 2016 were 23.8
30 Mt CO₂e, only some of which could be attributed to insurance as a driver.

31 Table 6.20 summarises the mitigation potentials for risk management options, with confidence
32 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
33 exhaustive) references upon which the evidence is based.

34

Table 6.20 Mitigation effects of response options based on risk management

| Integrated response option | Potential | Confidence | Citation |
|----------------------------|--|----------------|-------------------------------|
| Management of urban sprawl | No global estimates | No evidence | |
| Livelihood diversification | No global estimates | No evidence | |
| Use of local seeds | No global estimates | No evidence | |
| Disaster risk management | No global estimates | No evidence | |
| Risk sharing instruments | ->0.024 GtCO ₂ e yr ⁻¹ for crop insurance; likely some benefits for other risk sharing instruments | Low confidence | Claussen et al 2011; EPA 2018 |

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46

6.3.2 Potential of the integrated response options for delivering adaptation

In this section, the impacts of integrated response options on climate change adaptation are assessed.

6.3.2.1 Integrated response options based on land management

In this section, the impacts on climate change adaptation of integrated response options based on land management are assessed.

6.3.2.1.1 Integrated response options based on land management in agriculture

Increasing food productivity by practices such as sustainable intensification improves farm incomes and allows households to build assets for use in times of stress, thereby improving resilience (Campbell et al. 2014). By reducing pressure on land and increasing food production, increased food productivity could be beneficial for adaptation (Chapter 2; Section 6.3; Campbell et al. 2014). Pretty et al. (2018) report that 163 million farms occupying 4.53 Mkm² have passed a redesign threshold for application of sustainable intensification, suggesting the minimum number of people benefiting from increased productivity and adaptation benefits under sustainable intensification is >163 million, with the total likely to be far higher (Table 6.21).

Improved cropland management is a key climate adaptation option, potentially affecting more than 25 million people, including a wide range of technological decisions by farmers. Actions towards adaptation fall into two broad overlapping areas: (1) accelerated adaptation to progressive climate change over decadal time scales, for example integrated packages of technology, agronomy and policy options for farmers and food systems, including changing planting dates and zones, tillage systems, crop types and varieties, and (2) better management of agricultural risks associated with increasing climate variability and extreme events, for example improved climate information services and safety nets (Vermeulen et al. 2012b; Challinor et al. 2014; Lipper et al. 2014; Lobell 2014). In the same way, improved livestock management is another technological adaptation option potentially benefiting 1–25 million people. Crop and animal diversification are considered the most promising adaptation measures (Porter et al. 2014; Rojas-Downing et al. 2017a). In grasslands and rangelands, regulation of stocking rates, grazing field dimensions, establishment of exclosures and locations of drinking fountains and feeders are strategic decisions by farmers to improve grazing management (Taboada et al. 2011; Mekuria and Aynekulu 2013; Porter et al. 2014).

Around 30% of the world's rural population use trees across 46% of all agricultural landscapes (Lasco et al. 2014), meaning that up to 2.3 billion people benefit from agroforestry, globally (Table 6.21).

Agricultural diversification is key to achieve climatic resilience (Campbell et al. 2014; Cohn et al. 2017). Crop diversification is one important adaptation option to progressive climate change (Vermeulen et al. 2012) and it can improve resilience by engendering a greater ability to suppress pest outbreaks and dampen pathogen transmission, as well as by buffering crop production from the effects of greater climate variability and extreme events (Lin 2011).

Reduced conversion of grassland to cropland may lead to adaptation benefits by stabilising soils in the face of extreme climatic events (Lal 2001b), thereby increasing resilience, but since it would likely have a negative impact on food production / security (since croplands produce more food per unit area than grasslands), the wider adaptation impacts would likely be negative. However, there is no literature quantifying the global impact of avoidance of conversion of grassland to cropland on adaptation.

Integrated water management provides large co-benefits for adaptation (Dillon and Arshad 2016) by improving the resilience of food crop production systems to future climate change (Chapter 2; Table 6.7; Porter et al. 2014). Improving irrigation systems and integrated water resource management, such as enhancing urban and rural water supplies and reducing water evaporation losses (Dillon and

Arshad 2016), are significant options for enhancing climate adaptation. Many technical innovations (e.g., precision water management) can lead to beneficial adaptation outcomes by increasing water availability and the reliability of agricultural production, using different techniques of water harvesting, storage, and its judicious utilisation through farm ponds, dams, and community tanks in rainfed agriculture areas. Integrated water management response options that use freshwater would be expected to have few adverse side effects in regions where water is plentiful, but large adverse side effects in regions where water is scarce (Grey and Sadoff 2007; Liu et al. 2017; Scott et al. 2011).

Table 6.21 summarises the potentials for adaptation for agricultural response options, with confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not exhaustive) references upon which the evidence is based.

Table 6.21 Adaptation effects of response options based on land management in agriculture

| Integrated response option | Potential | Confidence | Citation |
|--|---------------------|-------------------|--|
| Increased food productivity | >163 million people | Medium confidence | Pretty et al. 2018 |
| Improved cropland management | >25 million people | Low confidence | Challinor et al. 2014; Lipper et al. 2014; Lobell 2014; Vermeulen et al. 2012b |
| Improved grazing land management | 1-25 million people | Low confidence | Porter et al. 2014 |
| Improved livestock management | 1-25 million people | Low confidence | Porter et al. 2014; Rojas-Downing et al. 2017 |
| Agroforestry | 2300 million people | Medium confidence | Lasco et al. 2014 |
| Agricultural diversification | >25 million people | Low confidence | Campbell et al. 2014; Cohn et al. 2017; Vermeulen et al. 2012b |
| Reduced grassland conversion to cropland | No global estimates | No evidence | |
| Integrated water management | 250 million people | Low confidence | Dillon and Arshad 2016; Liu et al. 2017 |

6.3.2.1.2 Integrated response options based on land management in forestry

Improved forest management positively impacts adaptation through limiting the negative effects associated with pollution (of air and fresh water), infections and other diseases, exposure to extreme weather events and natural disasters, and poverty (e.g., Smith et al. 2014c). There is high agreement on the fact that reduced deforestation and forest degradation positively impact adaptation and resilience of coupled human-natural systems. Based on the number of people affected by natural disasters (CRED 2015), the number of people depending to varying degrees on forests for their livelihoods (World Bank et al. 2009) and the current deforestation rate (Keenan et al. 2015), the estimated global potential effect for adaptation is large positive for improved forest management, and moderate positive for reduced deforestation when cumulated till the end of the century (Table 6.22). The uncertainty of these global estimates is high, e.g. the impact of reduced deforestation may be higher when the large biophysical impacts on the water cycle (and thus drought) from deforestation (e.g. Alkama & Cescatti 2016, etc) are taken into account (see Chapter 2).

More robust qualitative and some quantitative estimates are available at local and regional level. According to Karjalainen et al. (2009), reducing deforestation and habitat alteration contributes to limiting infectious diseases such as malaria in Africa, Asia, and Latin America, thus lowering the expenses associated with healthcare treatments. Bhattacharjee and Behera (2017) found that human lives lost due to floods increase with reducing forest cover and increasing deforestation rates in India. In addition, maintaining forest cover in urban contexts reduces air pollution and therefore avoids mortality of about one person per year per city in US, and up to 7.6 people per year in New York City

1 (Nowak et al. 2014). There is also evidence that reducing deforestation and degradation in mangrove
 2 plantations potentially improves soil stabilisation, and attenuates the impact of tropical cyclones and
 3 typhoons along the coastal areas in South and Southeast Asia (Chow 2018). At local scale, co-benefits
 4 between REDD+ and adaptation of local communities can potentially be substantial (Long 2013;
 5 Morita & Matsumoto 2017), even if often difficult to quantify, and not explicitly acknowledged
 6 (McElwee et al. 2017b).

7 Forest restoration may facilitate the adaptation and resilience of forests to climate change by
 8 enhancing connectivity between forest areas and conserving biodiversity hotspots (Locatelli et al.
 9 2011, 2015c; Ellison et al. 2017; Dooley and Kartha 2018b). Furthermore, forest restoration may
 10 improve ecosystem functionality and services, provide microclimatic regulation for people and crops,
 11 wood and fodder as safety nets, soil erosion protection and soil fertility enhancement for agricultural
 12 resilience, coastal area protection, water and flood regulation (Locatelli et al. 2015c).

13 Afforestation and reforestation are important climate change adaptation response options (Reyer et al.
 14 2009; Ellison et al. 2017a; Locatelli et al. 2015c), and can potentially help a large proportion of the
 15 global population to adapt to climate change and to associated natural disasters (Table 6.22). For
 16 example, trees general mitigate summer mean warming and temperature extremes (Findell et al. 2017;
 17 Sonntag et al. 2016).

18 Table 6.22 summarises the potentials for adaptation for forest response options, with confidence
 19 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 20 exhaustive) references upon which the evidence is based.

21 **Table 6.22 Adaptation effects of response options based on land management in forests**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|---------------------|-------------------|--|
| Improved forest management | > 25 million people | Low confidence | CRED 2015; World Bank et al. 2009 |
| Reduced deforestation and degradation | 1-25 million people | Low confidence | CRED 2015; Keenan et al. 2015; World Bank et al. 2009. The estimates consider a cumulated effect till the end of the century. |
| Reforestation and forest restoration | See afforestation | | |
| Afforestation | > 25 million people | Medium confidence | Griscom et al. 2017a; Reyer et al. 2009; Smith et al. 2014b; Sonntag et al. 2016. CRED 2015; World Bank, FAO, and IFAD, 2009. The estimates consider a cumulated effect till the end of the century. |

22

23 **6.3.2.1.3 Integrated response options based on land management of soils**

24 Soil organic carbon increase is promoted as an action for climate change adaptation. Since increasing
 25 soil organic matter content is a measure to address land degradation (see Section 6.2.1), and restoring
 26 degraded land helps to improve resilience to climate change, soil carbon increase is an important
 27 option for climate change adaptation. With around 120 thousand km² lost to degradation every year,
 28 and over 3.2 billion people negatively impacted by land degradation globally (IPBES 2018), practices
 29 designed to increase soil organic carbon have a large potential to address adaptation challenges (Table
 30 6.23).

31 Since soil erosion control prevents land degradation and desertification, it improves the resilience of
 32 agriculture to climate change and increases food production (Lal 1998; IPBES 2018), though the
 33 global number of people benefiting from improved resilience to climate change has not been reported

1 in the literature. Using figures from (FAO et al. 2015), Scholes et al. (2018) estimates that land losses
 2 due to erosion are equivalent to 1.5 Mkm² of land used for crop production to 2050, or 45 thousand
 3 km² yr⁻¹ (Foley et al. 2011). Control of soil erosion (water and wind) could benefit 11 Mkm² of
 4 degraded land (Lal 2014), and improve the resilience of at least some of the 3.2 billion people affected
 5 by land degradation (IPBES 2018), suggesting positive impacts on adaptation. Management of
 6 erosion is an important climate change adaptation measure, since it reduces the vulnerability of soils
 7 to loss under climate extremes, thereby increasing resilience to climate change (Garbrecht et al. 2015).

8 Prevention and/or reversion of topsoil salinisation may require a combined management of
 9 groundwater, irrigation techniques, drainage, mulching and vegetation, with all of these considered
 10 relevant for adaptation (Qadir et al. 2013; UNCTAD 2011; Dagar et al. 2016b). Taking into account
 11 the widespread diffusion of salinity problems, many people can benefit from its implementation by
 12 farmers. The relation between compaction prevention and/or reversion and climate adaption is less
 13 evident, and can be related to better hydrological soil functioning (Chamen et al. 2015; Epron et al.
 14 2016; Tullberg et al. 2018b).

15 Biochar has potential to benefit climate adaptation by improving the resilience of food crop
 16 production systems to future climate change by increasing yield in some regions and improving water
 17 holding capacity (Chapter 2; Section 6.4; Woolf et al. 2010; Sohi 2012). By increasing yield by 25%
 18 in the tropics (Jeffery et al. 2017), this could increase food production for 3.2 billion people affected
 19 by land degradation (IPBES 2018), thereby potentially improving their resilience to climate change
 20 shocks (Table 6.23). A requirement for large areas of land to provide feedstock for biochar could
 21 adversely impact adaptation, though the impact has not been quantified globally.

22 Table 6.23 summarises the potentials for adaptation for soil-based response options, with confidence
 23 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 24 exhaustive) references upon which the evidence is based.

25 **Table 6.23 Adaptation effects of response options based on land management of soils**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|---|----------------|---|
| Increased soil organic carbon content | Up to 3200 million people | Low confidence | Scholes et al. 2018 |
| Reduced soil erosion | Up to 3200 million people | Low confidence | Scholes et al. 2018 |
| Reduced soil salinisation | 1-25 million people | Low confidence | Dagar et al. 2016b; Qadir et al. 2013b; UNCTAD 2011 |
| Reduced soil compaction | <1 million people | Low confidence | Chamen et al. 2015c; Epron et al. 2016; Tullberg et al. 2018b |
| Biochar addition to soil | Up to 3200 million people; but potential negative (unquantified) impacts from land required from feedstocks | Low confidence | Jeffery et al. 2017 |

26 **6.3.2.1.4 Integrated response options based on land management across all/other ecosystems**

27 For fire management, Doerr et al. (2016) showed the number of people killed by wildfire was 1940,
 28 and the total number of people affected was 5.8 million from 1984 to 2013, globally. Johnston et al.
 29 (2012) showed the average mortality attributable to landscape fire smoke exposure was 339 thousand
 30 deaths annually. The regions most affected were sub-Saharan Africa (157 thousand) and Southeast
 31 Asia (110 thousand). Estimated annual mortality during La Niña was 262 thousand, compared with
 32 around 100 thousand excess deaths across Indonesia, Malaysia and Singapore (Table 6.24).
 33

34 Management of landslides and natural hazards are usually listed among planned adaptation options in
 35 mountainous and sloped hilly areas, where uncontrolled runoff and avalanches may cause climatic

1 disasters, affecting millions of people from both urban and rural areas. Landslide control requires both
2 increasing plant cover and engineering practices (see Table 6.8).

3 For management of pollution, including acidification, Anenberg et al. (2012) estimated that, for
4 PM_{2.5} and ozone, respectively, fully implementing reduction measures could reduce global
5 population-weighted average surface concentrations by 23–34% and 7–17% and avoid 0.6–4.4 and
6 0.04–0.52 million annual premature deaths globally in 2030. UNEP and WMO (2011) considered
7 emission control measures to reduce ozone and black carbon (BC) and estimated that 2.4 million
8 annual premature deaths (with a range of 0.7 to 4.6 million) from outdoor air pollution could be
9 avoided. West et al. (2013) estimated global GHG mitigation brings co-benefits for air quality and
10 would avoid 0.5±0.2, 1.3±0.5, and 2.2±0.8 million premature deaths in 2030, 2050, and 2100,
11 respectively.

12 There are no global data on the impacts of management of invasive species / encroachment on
13 adaptation.

14 Coastal wetlands provide a natural defence against coastal flooding and storm surges by dissipating
15 wave energy, reducing erosion and by helping to stabilise shore sediments, so restoration may provide
16 significant benefits for adaptation. The Ramsar Convention on Wetlands covers 1.5 Mkm² across
17 1674 sites Keddy et al. (2009). Coastal floods currently affect 93–310 million people (in 2010)
18 globally, and this could rise to 600 million people in 2100 with sea level rise, unless adaptation
19 measures are taken (Hinkel et al. 2014). The proportion of the flood-prone population that could avoid
20 these impacts through restoration of coastal wetlands has not been quantified, but this sets an upper
21 limit.

22 Avoided peat impacts and peatland restoration can help to regulate water flow and prevent
23 downstream flooding (Munang et al. 2014), but the global potential (in terms of number of people
24 who could avoid flooding through peatland restoration) has not been quantified.

25 There are no global estimates about the potential of biodiversity conservation to improve the
26 adaptation and resilience of local communities to climate change, in terms of reducing the number of
27 people affected by natural disasters. Nevertheless, it is widely recognised that biodiversity, ecosystem
28 health and resilience improves the adaptation potential (Jones et al. 2012). For example, tree species
29 mixture improves the resistance of stands to natural disturbances, such as drought, fires, and
30 windstorms (Jactel et al. 2017), as well as stability against landslides (Kobayashi and Mori 2017).
31 Moreover, Protected Areas play a key role for improving adaptation (Watson et al. 2014; Lopoukhine
32 et al. 2012), through reducing water flow, stabilising rock movements, creating physical barriers to
33 coastal erosion, improving resistance to fires, and buffering storm damages (Dudley et al. 2010).
34 33 out of 105 of the largest urban areas worldwide rely on protected areas for some, or all, of their drinking
35 water (Secretariat of the Convention on Biological Diversity 2008), indicating that many millions are
36 likely benefit from conservation practices.

37 Table 6.24 summarises the potentials for adaptation for soil-based response options, with confidence
38 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
39 exhaustive) references upon which the evidence is based.

40 **Table 6.24 Adaptation effects of response options based on land management of soils**

| Integrated response option | Potential | Confidence | Citation |
|--|--|-------------------|---|
| Fire management | > 5.8 million people affected by wildfire; max. 0.5 million deaths per year by smoke | Medium confidence | Doerr and Santín 2016; Johnston et al. 2012; Shannon et al., 2016 |
| Reduced landslides and natural hazards | >25 million people | Low confidence | Arnáez J et al. 2015; Gariano and Guzzetti 2016 |
| Reduced pollution | Prevent 0.5–4.6 million annual | Medium | Anenberg et al. 2012; |

| | | | |
|--|-----------------------------|----------------|--|
| including acidification | premature deaths globally | confidence | Shindell et al.; West et al. 2013; UNEP & WMO, 2011; |
| Management of invasive species / encroachment | No global estimates | No evidence | |
| Restoration and reduced conversion of coastal wetlands | up to 93-310 million people | Low confidence | Hinkel et al. 2014 |
| Restoration and reduced conversion of peatlands | No global estimates | No evidence | |
| Biodiversity conservation | Likely many millions | Low confidence | CBD, 2008 |

1

2 **6.3.2.1.5 Integrated response options based on land management specifically for CDR**

3 Enhanced weathering of minerals has been proposed as a mechanism of improving soil health and
4 food security (Beerling et al. 2018), but there is no literature estimating the global adaptation benefits.

5 Large-scale bioenergy and BECCS can require substantial amounts of cropland (Popp et al. 2017;
6 Calvin et al. 2014a; Smith et al. 2016c), forestland (Baker et al. 2019b; Favero and Mendelsohn
7 2017), and water (Chaturvedi et al. 2013; Smith et al. 2016; Fuss et al. 2018; Popp et al. 2011; Hejazi
8 et al. 2015b) suggesting that bioenergy and BECCS could have adverse side-effects for adaptation. In
9 some contexts, e.g., low inputs of fossil fuels and chemicals, limited irrigation, heat/drought tolerant
10 species, and using marginal land, bioenergy can have co-benefits for adaptation (Dasgupta et al. 2014;
11 Noble et al. 2014). However, no studies were found that quantify the magnitude of the effect.

12 Table 6.25 summarises the impacts on adaptation of land management response options specifically
13 for CDR, with confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6,
14 and indicative (not exhaustive) references upon which the evidence is based.

15 **Table 6.25 Adaptation effects of response options based on land management specifically for CDR**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------------|---|-------------------|--|
| Enhanced weathering of minerals | No global estimates | No evidence | |
| Bioenergy and BECCS | Potentially large negative consequences | Low confidence | Fuss et al. 2018b; Muller et al. 2017b; Smith et al. |

16

17 **6.3.2.2 Integrated response options based on value chain management**

18 In this section, the impacts on climate change adaptation of integrated response options based on
19 value chain management are assessed.

20 **6.3.2.2.1 Integrated response options based on value chain management through demand 21 management**

22 Decreases in pressure on land and decreases in production intensity associated with sustainable
23 healthy diets or reduced food waste could also benefit adaptation; however, the size of this effect is
24 not well quantified (Muller et al. 2017b).

25 Reducing food waste losses can relieve pressure on the global freshwater resource, thereby aiding
26 adaptation. Food losses account for 215 km³ yr⁻¹ of freshwater resources, which (Kummu et al. 2012)
27 report to be about 12–15% of the global consumptive water use. Given that 35% of the global
28 population is living under high water stress or shortage (Kummu et al. 2010), reducing food waste
29 could benefit 320–400 million people (12–15% of the 2681 million people affected by water stress /
30 shortage).

1 While no studies report quantitative estimates of the effect of material substitution on adaptation, the
 2 effects are expected to be similar to reforestation and afforestation if the amount of material
 3 substitution leads to an increase in forest area. Additionally, some studies indicate that wooden
 4 buildings, if properly constructed, could reduce fire risk compared to steel, which softens when
 5 burned (Gustavsson et al. 2006; Ramage et al. 2017a).

6 Table 6.26 summarises the impacts on adaptation of demand management options, with confidence
 7 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 8 exhaustive) references upon which the evidence is based.

9 **Table 6.26 Adaptation effects of response options based on demand management**

| Integrated response option | Potential | Confidence | Citation |
|---|------------------------|-------------------|---------------------|
| Dietary change | No global estimates | No evidence | Muller et al. 2017b |
| Reduced post-harvest losses | 320-400 million people | Medium confidence | Kummu et al. 2012a |
| Reduced food waste (consumer or retailer) | No global estimates | No evidence | Muller et al. 2017b |
| Material substitution | No global estimates | No evidence | |

10
 11 **6.3.2.2.2 Integrated response options based on value chain management through supply
 12 management**

13 It is estimated that 500 million smallholder farmers depend on agricultural businesses in developing
 14 countries (World Bank, 2017), meaning that better promotion of value-added products and improved
 15 efficiency and sustainability of food processing and retailing could potentially help up to 500 million
 16 people to adapt to climate change. However, figures on how sustainable sourcing in general could
 17 help farmers and forest management is mostly unquantified. More than 1 million farmers have
 18 currently been certified through various schemes (Tayleur et al. 2017), but how much this has helped
 19 them prepare for adaptation is unknown.

20 Management of supply chains has the potential to reduce vulnerability to price volatility. Consumers
 21 in lower income countries are most affected by price volatility, with sub-Saharan Africa and South
 22 Asia at highest risk (Regmi and Meade 2013; Fujimori et al. 2018a). However, understanding of the
 23 stability of food supply is one of the weakest links in global food system research (Wheeler and von
 24 Braun 2013) as instability is driven by a confluence of factors (Headey and Fan 2008). Food price
 25 spikes in 2007 increased the number of people under the poverty line by between 100 million people
 26 (Ivanic and Martin 2008) and 450 million people (Brinkman et al. 2009), and caused welfare losses of
 27 3% or more for poor households in many countries (Zezza et al. 2009). Food price stabilisation by
 28 China, India and Indonesia alone in 2007/2008 led to reduced staple food price for 2 billion people
 29 (Timmer 2009). Presumably, spending less on food frees up money for other activities, including
 30 adaptation, but it is unknown how much (Zezza et al. 2009; Ziervogel and Ericksen 2010). One
 31 example of reduction in staple food price costs to consumers in Bangladesh from food stability
 32 policies saved rural households USD887 million total (Torlesse et al. 2003b). Food supply stability
 33 through improved supply chains also potentially reduces conflicts (by avoiding food price riots, which
 34 occurred in countries with over 100 million total in population in 2007/2008), and thus increases
 35 adaptation capacity (Raleigh et al. 2015a).

36 There are no global estimates of the contribution of improved food transport and distribution, or of
 37 urban food systems, in contributing to adaptation, but since the urban population in 2018 was 4.2
 38 billion people, this sets the upper limit on those that could benefit.

39 Given that 65% (760 million) of poor working adults make a living through agriculture, increased
 40 energy efficiency in agriculture could benefit this 760 million people.

1 Table 6.27 summarises the impacts on adaptation of supply management options, with confidence
 2 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 3 exhaustive) references upon which the evidence is based.

4 **Table 6.27 Adaptation effects of response options based on demand management**

| Integrated response option | Potential | Confidence | Citation |
|--|---------------------|-------------------|--|
| Sustainable sourcing | > 1 million | Low confidence | Tayleur et al. 2017 |
| Management of supply chains | >100 million | Medium confidence | Ivanic and Martin 2008; Timmer 2009; Vermeulen et al. 2012b; Campbell et al. 2016; |
| Enhanced urban food systems | No global estimates | No evidence | |
| Improved food processing and retailing | 500 million people | Low confidence | World Bank 2017 |
| Improved energy use in food systems | 760 million | Low confidence | World Bank 2017 |

5

6 **6.3.2.3 Integrated response options based on risk management**

7 In this section, the impacts on climate change adaptation of integrated response options based on risk
 8 management are assessed.

9 Reducing urban sprawl is likely to provide adaptation co-benefits *via* improved human health
 10 (Frumkin 2002; Anderson 2017), as sprawl contributes to reduced physical activity, worse air
 11 pollution, and exacerbation of urban heat island effects and extreme heat waves (Stone et al. 2010).
 12 The most sprawling cities in the US have experienced extreme heat waves more than double those of
 13 denser cities, and “urban albedo and vegetation enhancement strategies have significant potential to
 14 reduce heat-related health impacts” (Stone et al. 2010). Other adaption co-benefits are less well
 15 understood. There are likely to be cost savings from managing planning growth (one study found 2%
 16 savings in metropolitan budgets, which can be then spent on adaptation planning) (Deal and Schunk
 17 2004).

18 Diversification is a major adaptation strategy and form of risk management, as it can help households
 19 smooth out income fluctuations and provide a broader range of options for the future (Osbaahr et al.
 20 2008; Adger et al. 2011; Thornton and Herrero 2014). Surveys of farmers in climate variable areas
 21 find that livelihood diversification is increasingly favoured as an adaptation option (Bryan et al.
 22 2013), although it is not always successful, since it can increase exposure to climate variability (Adger
 23 et al. 2011). There are over 570 million small farms in the world (Lowder et al. 2016), and many
 24 millions of smallholder agriculturalists already practice livelihood diversification by engaging in
 25 multiple forms of off-farm income (Rigg 2006). It is not clear, however, how many farmers have not
 26 yet practiced diversification and thus how many would be helped by supporting this response option.

27 Currently, millions of farmers still rely to some degree on local seeds. Use of local seeds can facilitate
 28 adaptation for many smallholders, as moving to use of commercial seeds can increase costs for
 29 farmers (Howard 2015). Seed networks and banks protect local agrobiodiversity and landraces, which
 30 are important to facilitate adaptation, as local landraces may be resilient to some forms of climate
 31 change (Coomes et al. 2015a; van Niekerk and Wynberg 2017a; Vasconcelos et al. 2013).

32 Disaster risk management is an essential part of adaptation strategies. The Famine Early Warning
 33 System funded by the USAID has operated across 3 continents since the 1980s, and many millions of
 34 people across 34 countries have access to early information on drought. Such information can assist
 35 communities and households in adapting to onset conditions (Hillbruner and Moloney 2012).

1 However, concerns have been raised as to how many people are actually reached by disaster risk
2 management and early warning systems; for example, less than 50% of respondents in Bangladesh
3 had heard a cyclone warning before it hit, even though an early warning system existed (Mahmud and
4 Prowse 2012). Further, there are concerns that current early warning systems “tend to focus on
5 response and recovery rather than on addressing livelihood issues as part of the process of reducing
6 underlying risk factors,” (Birkmann et al. 2015a), leading to less adaptation potential being realised.

7 Local risk sharing instruments like rotating credit or loan groups can help buffer farmers against
8 climate impacts and help facilitate adaptation. Both index and commercial crop insurance offers some
9 potential for adaptation, as it provides a means of buffering and transferring weather risk, saving
10 farmers the cost of crop losses (Meze-Hausken et al. 2009; Patt et al. 2010). However, overly
11 subsidised insurance can undermine the market’s role in pricing risks and thus depress more rapid
12 adaptation strategies (Skees and Collier 2012; Jaworski 2016) and increase the riskiness of decision-
13 making (McLeman and Smit 2006). For example, availability of crop insurance was observed to
14 reduce farm-level diversification in the US, a factor cited as increasing adaptive capacity (Sanderson
15 et al. 2013b) and crop insurance-holding soybean farmers in the US have been less likely to adapt to
16 extreme weather events than those not holding insurance (Annan and Schlenker 2015). It is unclear
17 how many people worldwide use insurance as an adaptation strategy; (Platteau et al. 2017) suggest
18 less than 30% of smallholders take out any form of insurance), but it is likely in the millions.

19 Table 6.28 summarises the impacts on adaptation of risk management options, with confidence
20 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
21 exhaustive) references upon which the evidence is based.

22 **Table 6.28 Adaptation effects of response options based on risk management**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------------|---|-------------------|------------------------------|
| Management of urban sprawl | Unquantified but likely to be many millions | Low confidence | Stone et al. 2010 |
| Livelihood diversification | >100 million likely | Low confidence | Morton 2007; Rigg 2006 |
| Use of local seeds | Unquantified but likely to be many millions | Low confidence | Louwaars 2002; Santilli 2012 |
| Disaster risk management | >100 million | High confidence | Hillbruner and Moloney 2012 |
| Risk sharing instruments | Unquantified but likely to be several million | Low confidence | Platteau et al. 2017 |

24 **6.3.3 Potential of the integrated response options for addressing desertification**

25 In this section, the impacts of integrated response options on desertification are assessed.

26 **6.3.3.1 Integrated response options based on land management**

27 In this section, the impacts on desertification of integrated response options based on land
28 management are assessed.

29 **6.3.3.1.1 Integrated response options based on land management in agriculture**

30 Burney et al. (2010) estimated that an additional global cropland area of 11.11–15.14 Mkm² would
31 have been needed if productivity had not increased between 1961 and 2000. Given that agricultural
32 expansion is a main driver of desertification (FAO et al. 2015), increased food productivity could
33 have prevented up to 11.11–15.14 Mkm² from exploitation and desertification (Table 6.10).

34 Improved cropland, livestock and grazing land management are strategic options aiming at prevention
35 of desertification, and may include crop and animal selection, optimised stocking rates, changed

1 tillage and/or cover crops, to land use shifting from cropland to rangeland, in general targeting
 2 increases in ground cover by vegetation, and protection against wind erosion (Schwilch et al. 2014;
 3 Bestelmeyer et al. 2015). Considering the widespread distribution of deserts and desertified lands
 4 globally, more than 10 Mkm² could benefit from improved management techniques.

5 Agroforestry can help stabilise soils to prevent desertification (Section 6.3.2.1.1), so given that there
 6 are is around 10 Mkm² of land with more than 10% tree cover (Garrity, 2012), agroforestry could
 7 benefit up to 10 Mkm² of land.

8 Agricultural diversification to prevent desertification may include the use of crops with manures,
 9 legumes, fodder legumes and cover crops combined with conservation tillage systems (Schwilch et al.
 10 2014). These practices can be considered to be part of improved crop management options (see
 11 above) and aim at increasing ground coverage by vegetation and controlling wind erosion losses.

12 Since shifting from grassland to the annual cultivation of crops increases erosion and soil loss, there
 13 are significant benefits for desertification control, by stabilising soils in arid areas (Chapter 3).
 14 Cropland expansion during 1985 to 2005 was 359 thousand km², or 17.4 thousand km² yr⁻¹ (Foley et
 15 al. 2011). Not all of this expansion will be from grasslands or in desertified areas, but this value sets
 16 the maximum contribution of prevention of conversion of grasslands to croplands, a small global
 17 benefit for desertification control (Table 6.10).

18 Integrated water management strategies such as water-use efficiency and irrigation, improve soil
 19 health through increase in soil organic matter content, thereby delivering benefits for prevention or
 20 reversal of desertification (Chapter 3; Baumhardt et al. 2015; Datta et al. 2000; Evans and Sadler
 21 2008; He et al. 2015). Climate change will amplify existing stress on water availability and on
 22 agricultural systems, particularly in semi-arid environments (AR5; Chapter 3). In 2011, semiarid
 23 ecosystems in the southern hemisphere contributed 51% of the global net carbon sink (Poulter et al.,
 24 2014). These results suggest that arid ecosystems could be an important global carbon sink, depending
 25 on soil water availability.

26 Table 6.29 summarises the impacts on desertification of agricultural options, with confidence
 27 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 28 exhaustive) references upon which the evidence is based.

29 **Table 6.29 Effects on desertification of response options in agriculture**

| Integrated response option | Potential | Confidence | Citation |
|--|--|-------------------|---|
| Increased food productivity | 11.1-15.1 Mkm ² | Low confidence | Burney et al. 2010 |
| Improved cropland management | 10 Mkm ² | Low confidence | Schwilch et al. 2014 |
| Improved grazing land management | 0.5-3 Mkm ² | Low confidence | Schwilch et al. 2014 |
| Improved livestock management | 0.5-3 Mkm ² | Low confidence | Miao et al. 2015; Squires and Karami 2005 |
| Agroforestry | 10 Mkm ² (with >10% tree cover) | Medium confidence | Garrity (2012) |
| Agricultural diversification | 0.5-3 Mkm ² | Low confidence | Lambin and Meyfroidt 2011; Schwilch et al. 2014 |
| Reduced grassland conversion to cropland | up to 17.4 thousand km ² yr ⁻¹ | Low confidence | Foley et al. 2011 |
| Integrated water management | 10 thousand km ² | Low confidence | Pierzynski et al., 2017; UNCCD, 2011 |

30 **6.3.3.1.2 Integrated response options based on land management in forestry**

31 Forests are important to help to stabilise land and regulate water and microclimate (Locatelli et al.
 32 2015c). Based on the extent of dry forest at risk of desertification (Núñez et al. 2010; Bastin et al.
 33

1 2017), the estimated global potential effect for avoided desertification is large for both improved
 2 forest management and for reduced deforestation and forest degradation when cumulated for at least
 3 20 years (Table 6.30). The uncertainty of these global estimates is high. More robust qualitative and
 4 some quantitative estimates are available at regional level. For example, it has been simulated that
 5 human activity (i.e., land management) contributed to 26% of the total land reverted from
 6 desertification in Northern China between 1981 and 2010 (Xu et al. 2018). In Thailand, it was found
 7 that the desertification risk is reduced when the land use is changed from bare lands to agricultural
 8 lands and forests, and from non-forests to forests; conversely, the desertification risk increases when
 9 converting forests and denuded forests to bare lands (Wijitkosum 2016).

10 Afforestation, reforestation and forest restoration are land management response options that are used
 11 to prevent desertification. Forests tend to maintain water and soil quality by reducing runoff and
 12 trapping sediments and nutrients (Idris Medugu et al. 2010a; Salvati et al. 2014a), but planting of non-
 13 native species in semi-arid regions can deplete soil water resources if they have high
 14 evapotranspiration rates (Feng et al.; Yang et al.). Afforestation and reforestation programmes can be
 15 deployed over large areas of the Earth, so can create synergies in areas prone to desertification. Global
 16 estimates of land potentially available for afforestation are up to 25.8 Mkm² by the end of the century,
 17 depending on a variety of assumptions on socio-economic developments and climate policies
 18 (Griscom et al. 2017; Kreidenweis et al. 2016a; Popp et al. 2017). The higher end of this range is
 19 achieved under the assumption of a globally uniform reward for carbon uptake in the terrestrial
 20 biosphere, and it is halved by considering tropical and subtropical areas only to minimise albedo
 21 feedbacks (Kreidenweis et al. 2016a). When safeguards are introduced (e.g., excluding existing
 22 cropland for food security, boreal areas, etc.), the area available declines to about 6.8 Mkm² (95%
 23 confidence interval of 2.3 and 11.25 Mkm²), of which about 4.72 Mkm² is in the tropics and 2.06
 24 Mkm² is in temperate regions (Griscom et al. 2017a; Table 6.30).

25 Table 6.30 summarises the impacts on desertification of forestry options, with confidence estimates
 26 based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not exhaustive)
 27 references upon which the evidence is based.

28 **Table 6.30 Effects on desertification of response options in forests**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|--|-------------------|--|
| Improved forest management | > 3 Mkm ² | Low confidence | Bastin et al. 2017; Núñez et al. 2010 |
| Reduced deforestation and degradation | > 3 Mkm ² (effects cumulated for at least 20 years) | Low confidence | Bastin et al. 2017; Keenan et al. 2015; Núñez et al. 2010 |
| Reforestation and forest restoration | See afforestation | | |
| Afforestation | 2-25.8 Mkm ² by the end of the century | Medium confidence | Griscom et al. 2017a; Kreidenweis et al. 2016a; Popp et al. 2017 |

30 **6.3.3.1.3 Integrated response options based on land management of soils**

31 With over 2.7 billion people affected globally by desertification (IPBES 2018), practices to increase
 32 soil organic carbon content are proposed as actions to address desertification, and could be applied to
 33 an estimated 11.37 Mkm² of desertified soils (Lal 2001a; Table 6.31).

34 Control of soil erosion could have large benefits for desertification control. Using figures from (FAO
 35 et al. 2015), Scholes et al. (2018) estimated that land losses due to erosion to 2050 are equivalent to
 36 1.5 Mkm² of land from crop production, or 45 thousand km² yr⁻¹ (Foley et al. 2011), so soil erosion
 37 control could benefit up to 1.50 Mkm² of land in the coming decades. Lal (2001a) estimated that
 38 desertification control (using soil erosion control as one intervention) could benefit 11.37 Mkm² of
 39 desertified land globally (Table 6.10).

1 Oldeman et al. (1991) estimated the global extent soil affected by salinisation is 0.77 Mkm² yr⁻¹,
2 which sets the upper limit on the area that could benefit from measures to address soil salinisation
3 (Table 6.31).

4 In degraded arid grasslands, shrublands and rangelands, desertification can be reversed by alleviation
5 of soil compaction through installation of enclosures and removal of domestic livestock (Allington et
6 al. 2010), but there are no global estimates of potential (Tale 6.31).

7 Biochar could potentially deliver benefits in efforts to address desertification though improving water
8 holding capacity (Woolf et al. 2010; Sohi 2012), but the global effect is not quantified.

9 Table 6.31 summarises the impacts on desertification of soil-based options, with confidence estimates
10 based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not exhaustive)
11 references upon which the evidence in based.

12 **Table 6.31 Effects on desertification of land management of soils**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|--|-------------------|---|
| Increased soil organic carbon content | Up to 11.37 Mkm ² | Medium confidence | Lal 2001a |
| Reduced soil erosion | Up to 11.37 Mkm ² | Medium confidence | Lal 2001a |
| Reduced soil salinisation | 0.77 Mkm ² yr ⁻¹ | Medium confidence | Oldeman et al. 1991 |
| Reduced soil compaction | No global estimates | No evidence | FAO and ITPS 2015; Hamza and Anderson 2005b |
| Biochar addition to soil | No global estimates | No evidence | |

13

14 **6.3.3.1.4 Integrated response options based on land management across all/other ecosystems**

15 For fire management, Arora and Melton (2018) estimated, using models and GFED4.1s0 data, that
16 burned area over the 1997–2014 period was 4.834–4.855 Mkm² yr⁻¹. Randerson et al. (2012) estimated
17 small fires increased total burned area globally by 35% from 3.45 to 4.64 Mkm² yr⁻¹ during the period
18 2001–2010. Tansey et al. (2004) estimated over 3.5 Mkm² yr⁻¹ of burned areas were detected in the
19 year 2000 (Table 6.32).

20 Although slope and slope aspect are predictive factors of desertification occurrence, the factors with
21 the greatest influence are land cover factors, such as normalised difference vegetation index (NDVI)
22 and rangeland classes (Djeddaoui et al. 2017). Therefore, prevention of landslides and natural hazards
23 exert indirect influence on the occurrence of desertification.

24 The global extent of chemical soil degradation (salinisation, pollution, and acidification) is about 1.03
25 Mkm² yr⁻¹ (Oldeman et al. 1991), giving the maximum extent of land that could benefit from the
26 management of pollution and acidification.

27 There are no global data on the impacts of management of invasive species / encroachment on
28 desertification, though the impact is presumed to be positive. There are no studies examining the
29 potential role of restoration and avoided conversion of coastal wetlands on desertification.

30 There are no impacts of peatland restoration for prevention of desertification, as peatlands occur in
31 wet areas and deserts in arid areas, so they are not connected.

32 For management of pollution, including acidification, Oldeman et al. (1991) estimated global extent
33 of chemical soil degradation, with 0.77 Mkm² yr⁻¹ affected by salinisation, 0.21 Mkm² yr⁻¹ affected by
34 pollution, and 0.06 Mkm² yr⁻¹ affected by pollution (total: 1.03 Mkm² yr⁻¹), so this is the area that
35 could potentially benefit from pollution management measures.

1 Biodiversity conservation measures can interact with desertification, but the literature contains no
2 global estimates of potential.

3 Table 6.32 summarises the impacts on desertification of options on all/other ecosystems, with
4 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
5 (not exhaustive) references upon which the evidence is based.

6 **Table 6.32 Effects on desertification of response options on all/other ecosystems**

| Integrated response option | Potential | Confidence | Citation |
|--|---|-------------------|--|
| Fire management | Up to 3.5-4.9 Mkm ² yr ⁻¹ | Medium confidence | Arora and Melton 2018; Randerson et al. 2012; Tansey et al. 2004 |
| Reduced landslides and natural hazards | >0 | Low confidence | Djeddaoui et al.; Noble et al. 2014 |
| Reduced pollution including acidification | 1.03 Mkm ² yr ⁻¹ | Low confidence | Oldeman et al. 1991 |
| Management of invasive species / encroachment | No global estimates | No evidence | |
| Restoration and reduced conversion of coastal wetlands | No global estimates | No evidence | |
| Restoration and reduced conversion of peatlands | No impact | | |
| Biodiversity conservation | No global estimates | No evidence | |

7

8 **6.3.3.1.5 Integrated response options based on land management specifically for CDR**

9 While spreading of crushed minerals onto land as part of enhanced weathering may provide soil /
10 plant nutrients in nutrient-depleted soils (Beerling et al. 2018), there is no literature reporting on the
11 potential global impacts of this in addressing desertification.

12 Large-scale production of bioenergy can require significant amounts of land (Smith et al. 2016d;
13 Clarke and Jiang 2014a; Popp et al. 2017), with as much as 15 Mkm² in 2100 in 2°C scenarios (Popp
14 et al. 2017), increasing pressures for desertification (Table 6.33).

15 Table 6.33 summarises the impacts on desertification of options specifically for CDR, with
16 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
17 (not exhaustive) references upon which the evidence is based.

18 **Table 6.33 Effects on desertification of response options specifically for CDR**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------|--|-------------------|---|
| Enhanced weathering of minerals | No global estimates | No evidence | |
| Bioenergy and BECCS | Negative impact on up to 15 Mkm ² | Medium confidence | Clarke et al. 2014a; Popp et al. 2017; Smith et al. |

19

20 **6.3.3.2 Integrated response options based on value chain management**

21 In this section, the impacts on desertification of integrated response options based on value chain
22 management are assessed.

23 **6.3.3.2.1 Integrated response options based on value chain management through demand 24 management**

25 Dietary change and waste reduction both result in decreased cropland and pasture extent (Bajželj et al.
26 2014a; Stehfest et al. 2009; Tilman and Clark 2014), reducing the pressure for desertification (Table
27 6.34).

1 Reduced post-harvest losses could spare 1.98 Mkm² of cropland globally (Kummu et al. 2012). Not
 2 all of this land could be subject to desertification pressure, so this represents that maximum area that
 3 could be relieved from desertification pressure by reduction of post-harvest losses. No studies were
 4 found linking material substitution to desertification.

5 Table 6.34 summarises the impacts on desertification of demand management options, with
 6 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
 7 (not exhaustive) references upon which the evidence is based.

8 **Table 6.34 Effects on desertification of response options based on demand management**

| Integrated response option | Potential | Confidence | Citation |
|---|-------------------------|-------------------|--|
| Dietary change | 0.80-5 Mkm ² | Low confidence | Alexander et al. 2016; Bajželj et al. 2014; Stehfest et al. 2009; Tilman and Clark 2014a |
| Reduced post-harvest losses | <1.98 Mkm ² | Low confidence | Kummu et al. 2012a |
| Reduced food waste (consumer or retailer) | 1.4 Mkm ² | Low confidence | Bajželj et al. 2014 |
| Material substitution | No global estimates | No evidence | |

9

10 **6.3.3.2 Integrated response options based on value chain management through supply** 11 **management**

12 There are no global estimates of the impact on desertification of sustainable sourcing, management of
 13 supply chains, enhanced urban food systems, improved food processing, or improved energy use in
 14 agriculture.

15 Table 6.35 summarises the impacts on desertification of supply management options, with confidence
 16 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 17 exhaustive) references upon which the evidence is based.

18 **Table 6.35 Effects on desertification of response options based on supply management**

| Integrated response option | Potential | Confidence | Citation |
|--|---------------------|-------------------|-----------------|
| Sustainable sourcing | No global estimates | No evidence | |
| Management of supply chains | No global estimates | No evidence | |
| Enhanced urban food systems | No global estimates | No evidence | |
| Improved food processing and retailing | No global estimates | No evidence | |
| Improved energy use in food systems | No global estimates | No evidence | |

19

20 **6.3.3.3 Integrated response options based on risk management**

21 In this section, the impacts on desertification of integrated response options based on risk
 22 management are assessed.

23 There are regional case studies of urban sprawl contributing to desertification in Mediterranean
 24 climates in particular (Barbero-Sierra et al. 2013b; Stellmes et al. 2013), but no global figures.

25 Diversification may deliver some benefits for addressing desertification when it involves greater use
 26 of tree crops that may reduce the need for tillage (Antwi-Agyei et al. 2014). Many anti-
 27 desertification programmes call for diversification (Stringer et al. 2009), but there is little
 28 evidence on how many households had done so (Herrmann and Hutchinson 2005). There are no
 29 numbers for global impacts.

30 The literature is unclear on whether the use of local seeds has any relationship to desertification,
 31 although some local seeds are likely more adapted to arid climates and less likely to degrade land than

1 commercial introduced varieties (Mousseau 2015). Some anti-desertification programmes have also
2 shown more success using local seed varieties (Bassoum and Ghiggi 2010; Nunes et al. 2016).

3 Some disaster risk management approaches can have impacts on reducing desertification, like the
4 Global Drought Early Warning System (GDEWS) (currently in development), which will monitor
5 precipitation, soil moisture, evapotranspiration, river flows, groundwater, agricultural productivity
6 and natural ecosystem health. It may have some potential co-benefits to reduce desertification (Pozzi
7 et al. 2013). However, there are no figures yet for how much land area will be covered by such early
8 warning systems.

9 Risk sharing instruments, like pooling labour or credit, could help communities invest in anti-
10 desertification actions, but evidence is missing. Commercial crop insurance is likely to deliver no co-
11 benefits for prevention and reversal of desertification, as evidence suggests that subsidised insurance,
12 in particular, can increase crop production in marginal lands. Crop insurance could have been
13 responsible for shifting up to 0.9% of rangelands to cropland in the Upper US Midwest (Claassen et
14 al. 2011).

15 Table 6.36 summarises the impact on desertification for options based on risk management, with
16 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
17 (not exhaustive) references upon which the evidence is based.

18 **Table 6.36 Effects on desertification of response options based on risk management**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------------|--|-------------------|------------------------------|
| Management of urban sprawl | >5 thousand km ² | Low confidence | Barbero-Sierra et al. 2013b |
| Livelihood diversification | No global estimates | Low confidence | Herrmann and Hutchinson 2005 |
| Use of local seeds | No global estimates | No evidence | |
| Disaster risk management | No global estimates | No evidence | Pozzi et al. 2013a |
| Risk sharing instruments | Likely negative impacts but not quantified | Low confidence | Claassen et al. 2011 |

19 **6.3.4 Potential of the integrated response options for addressing land degradation**

20 In this section, the impacts of integrated response options on land degradation are assessed.

21 **6.3.4.1 Integrated response options based on land management**

22 In this section, the impacts on land degradation of integrated response options based on land
23 management are assessed.

24 **6.3.4.1.1 Integrated response options based on land management in agriculture**

25 Burney et al. (2010) estimated that an additional global cropland area of 11.11–15.14 Mkm² would
26 have been needed if productivity had not increased between 1961 and 2000. As for desertification,
27 given that agricultural expansion is a main driver of land degradation (FAO and ITPS 2015),
28 increased food productivity has prevented up to 11.11–15.14 Mkm² from exploitation and land
29 degradation (Table 6.37).

30 Land degradation can be addressed by the implementation of improved cropland, livestock and
31 grazing land management practices, such as those outlined in the recently published Voluntary
32 Guidelines for Sustainable Soil Management (FAO 2017b). Each one could potentially affect
33 extensive surfaces, not less than 10 Mkm². The Guidelines include a list of practices aiming at
34 minimising soil erosion, enhancing soil organic matter content, fostering soil nutrient balance and
35 cycles, preventing, minimising and mitigating soil salinisation and alkalinisation, soil contamination,
36 soil acidification, and soil sealing, soil compaction, and improving soil water management. Land
37

1 cover and land cover change are key factors and indicators of land degradation. In many drylands,
 2 land cover is threatened by overgrazing, so management of stocking rate and grazing can help to
 3 prevent the advance of land degradation Smith et al. (2016c).

4 Agroforestry can help stabilise soils to prevent land degradation, so given that there are is around 10
 5 Mkm² of land with more than 10% tree cover (Garrity, 2012), agroforestry could benefit up to 10
 6 Mkm² of land.

7 Agricultural diversification usually aims at increasing climate and food security resilience, such as
 8 under “climate smart agriculture” approaches (Lipper et al. 2014). Both objectives are closely related
 9 to land degradation prevention, potentially affecting 1–5 Mkm².

10 Shifting from grassland to tilled crops increases erosion and soil loss, so there are significant benefits
 11 for addressing land degradation, by stabilising degraded soils (Chapter 3). Since cropland expansion
 12 during 1985 to 2005 was 17.4 thousand km² yr⁻¹ (Foley et al.,2009), and not all of this expansion will
 13 be from grasslands or degraded land, the maximum contribution of prevention of conversion of
 14 grasslands to croplands is 17.4 thousand km² yr⁻¹, a small global benefit for control of land
 15 degradation (Tale 6.37).

16 Most land degradation processes that are sensitive to climate change pressures (e.g. erosion, decline in
 17 soil organic matter, salinisation, waterlogging, drying of wet ecosystems) can benefit from integrated
 18 water management. Integrated water management options include management to reduce aquifer and
 19 surface water depletion, and to prevent over extraction, and provide direct co-benefits for prevention
 20 of land degradation. Land management practices implemented for climate change mitigation may also
 21 affect water resources. Globally, water erosion is estimated to result in the loss of 23–42 MtN and
 22 14.6–26.4 MtP annually (Pierzynski et al., 2017). Forests influence the storage and flow of water in
 23 watersheds (Eisenbies et al. 2007) and are therefore important for regulating how climate change will
 24 impact landscapes.

25 Table 6.37 summarises the impact on land degradation of options in agriculture, with confidence
 26 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 27 exhaustive) references upon which the evidence in based.

28 **Table 6.37 Effects on land degradation of response options in agriculture**

| Integrated response option | Potential | Confidence | Citation |
|--|--|-------------------|--------------------------------------|
| Increased food productivity | 11.11-15.14 Mkm ² | Medium confidence | Burney et al. 2010 |
| Improved cropland management | 10 Mkm ² | Low confidence | Lal 2015; Smith et al. 2016c |
| Improved grazing land management | 10 Mkm ² | Low confidence | Smith et al. 2016c |
| Improved livestock management | 10 Mkm ² | Low confidence | Lal 2015; Smith et al. 2016c |
| Agroforestry | 10 Mkm ² (with >10% tree cover) | Medium confidence | Garrity 2012 |
| Agricultural diversification | 1-5 Mkm ² | Medium confidence | Lambin and Meyfroidt 2011 |
| Reduced grassland conversion to cropland | Up to 17.4 thousand km ² yr ⁻¹ | Low confidence | Foley et al. 2011 |
| Integrated water management | 0.01 Mkm ² | Medium confidence | Pierzynski et al., 2017; UNCCD, 2011 |

29

30 **6.3.4.1.2 Integrated response options based on land management in forestry**

31 Based on the extent of forest exposed to degradation (Gibbs and Salmon 2015), the estimated global
 32 potential effect for reducing land degradation, e.g. through reduced soil erosion (Borrelli et al. 2017),
 33 is large for both improved forest management and for reduced deforestation and forest degradation

1 when cumulated for at least 20 years (Table 6.38) The uncertainty of these global estimates is high.
 2 More robust qualitative and some quantitative estimates are available at regional level. For example,
 3 in Indonesia, Santika et al. (2017) demonstrated that reduced deforestation (Sumatra and Kalimantan
 4 islands) contributed to reduce significantly land degradation.

5 Forest restoration is a key option to achieve the overarching frameworks to reduce land degradation at
 6 global scale, such as for example, Zero Net Land Degradation (ZNLD; UNCCD 2012) and Land
 7 Degradation Neutrality (LDN), not only in drylands (Safriel 2017). Indeed, it has been estimated that
 8 more than 20 Mkm² are suitable for forest and landscape restoration, of which 15 Mkm² may be
 9 devoted to mosaic restoration (UNCCD 2012). Moreover, the Bonn Challenge aims to restore 1.5
 10 Mkm² of deforested and degraded land by 2020, and 3.5 Mkm² by 2030
 11 (<http://www.bonnchallenge.org/content/challenge>). Under a restoration and protection scenario
 12 (implementing restoration targets), Wolff et al. (2018) simulated that there will be a global increase in
 13 net tree cover of about 4 Mkm² by 2050 (Table 6.38). At local level, Brazil's Atlantic Restoration
 14 Pact aims to restore 0.15 Mkm² of forest areas in 40 years (Melo et al. 2013). The Y Ikatu Xingu
 15 campaign (launched in 2004) aims to contain deforestation and degradation processes by reversing the
 16 liability of 3 thousand km² in the Xingu Basin, Brazil (Durigan et al. 2013).

17 Afforestation and reforestation are land management options frequently used to address land
 18 degradation (see Section 6.3.3.1.2 for details; Table 6.38).

19 Table 6.38 summarises the impact on land degradation of options in forestry, with confidence
 20 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 21 exhaustive) references upon which the evidence is based.

22 **Table 6.38 Effects on land degradation of response options in forestry**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|--|-------------------|--|
| Improved forest management | > 3 Mkm ² | Low confidence | Gibbs and Salmon 2015 |
| Reduced deforestation and degradation | > 3 Mkm ² (effects cumulated for at least 20 years) | Low confidence | Gibbs and Salmon 2015; Keenan et al. 2015 |
| Reforestation and forest restoration | 20 Mkm ² suitable for restoration > 3 Mkm ² by 2050 (net increase in tree cover for forest restoration) | Medium confidence | UNCCD 2012; Wolff et al. 2018 |
| Afforestation | 2-25.8 Mkm ² by the end of the century | Low confidence | Griscom et al. 2017a; Kreidenweis et al. 2016a; Popp et al. 2017 |

23

24 **6.3.4.1.3 Integrated response options based on land management of soils**

25 Increasing soil organic matter content is a measure to address land degradation. With around 120
 26 thousand km² lost to degradation every year, and over 3.2 billion people negatively impacted by land
 27 degradation globally (IPBES 2018), practices designed to increase soil organic carbon have a large
 28 potential to address land degradation, estimated to affect over 11 Mkm² globally (Lal, 2004; Table
 29 6.39).

30 Control of soil erosion could have large benefits for addressing land degradation. Soil erosion control
 31 could benefit up to 1.50 Mkm² of land to 2050 (IPBES 2018). Lal (2004) suggested interventions to
 32 prevent wind and water erosion (two of the four main interventions proposed to address land
 33 degradation), could restore 11 Mkm² of degraded and desertified soils globally (Table 6.39).

34 Oldeman et al. (1991) estimated the global extent soil affected by salinisation is 0.77 Mkm² yr⁻¹,
 35 which sets the upper limit on the area that could benefit from measures to address soil salinisation

1 (Table 6.39). The global extent of chemical soil degradation (salinisation, pollution, and acidification)
 2 is about 1.03 Mkm² (Oldeman et al. 1991) giving the maximum extent of land that could benefit from
 3 the management of pollution and acidification.

4 Biochar could provide moderate benefits for the prevention or reversal of land degradation, by
 5 improving water holding capacity, improving nutrient use efficiency, managing heavy metal pollution
 6 and other co-benefits (Sohi 2012), though the global effects are not quantified.

7 Table 6.39 summarises the impact on land degradation of soil-based options, with confidence
 8 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 9 exhaustive) references upon which the evidence is based.

10 **Table 6.39 Effects on land degradation of soil-based response options**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|--|-------------------|---|
| Increased soil organic carbon content | 11 Mkm ² | Medium confidence | Lal 2004 |
| Reduced soil erosion | 11 Mkm ² | Medium confidence | Lal 2004 |
| Reduced soil salinisation | 0.77 Mkm ² yr ⁻¹ | Medium confidence | Qadir et al. 2013a; FAO 2016; |
| Reduced soil compaction | 10 Mkm ² | Low confidence | FAO and ITPS 2015; Hamza and Anderson 2005a |
| Biochar addition to soil | Positive but not quantified globally | Low confidence | Chapter 4 |

11

12 **6.3.4.1.4 Integrated response options based on land management across all/other ecosystems**

13 For fire management, details of estimates of the impact of wildfires (and thereby the potential impact
 14 of their suppression) are given in Section 6.3.3.1.4 (Table 6.40).

15 Management of landslides and natural hazards aims at controlling a severe land degradation process
 16 affecting sloped and hilly areas, many of them with poor rural inhabitants (FAO et al. 2015; Gariano
 17 and Guzzetti 2016b), but the global potential has not been quantified.

18 There are no global data on the impacts of management of invasive species / encroachment on land
 19 degradation, though the impact is presumed to be positive.

20 Since large areas of coastal wetlands are degraded, restoration could potentially deliver moderate
 21 benefits for addressing land degradation, with 0.29 Mkm² globally considered feasible for restoration
 22 (Griscom et al. 2017a; Table 6.40).

23 Considering that large areas (0.46 Mkm²) of global peatlands are degraded and considered suitable for
 24 restoration (Griscom et al. 2017), peatland restoration could deliver moderate benefits for addressing
 25 land degradation (Table 6.40).

26 There are no global estimates of the effects of biodiversity conservation on reducing degraded lands.
 27 However, at local scale, biodiversity conservation programmes have been demonstrated to stimulate
 28 gain of forest cover over large areas over the last three decades (e.g. in China; Zhang et al. 2013).
 29 Management of wild animals can influence land degradation processes by grazing, trampling and
 30 compacting soil surfaces, thereby altering surface temperatures and chemical reactions affecting
 31 sediment and carbon retention (Cromsigt et al. 2018).

32 Table 6.40 summarises the impact on land degradation of options in all/other ecosystems, with
 33 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
 34 (not exhaustive) references upon which the evidence is based.

35 **Table 6.40 Effects on land degradation of response options in all/other ecosystems**

| Integrated response option | Potential | Confidence | Citation |
|--|---|-------------------|--|
| Fire management | Up to 3.5-4.9 Mkm ² yr ⁻¹ | Medium confidence | Arora and Melton 2018; Randerson et al. 2012; Tansey et al. 2004 |
| Reduced landslides and natural hazards | 1-5 Mkm ² | Low confidence | FAO and ITPS 2015; Gariano and Guzzetti 2016 |
| Reduced pollution including acidification | ~1.03 Mkm ² | Low confidence | Oldeman et al. 1991 |
| Management of invasive species / encroachment | No global estimates | No evidence | |
| Restoration and reduced conversion of coastal wetlands | 0.29 Mkm ² | Medium confidence | Griscom et al. 2017a |
| Restoration and reduced conversion of peatlands | 0.46 Mkm ² | Medium confidence | Griscom et al. 2017a |
| Biodiversity conservation | No global estimates | No evidence | |

1

2 **6.3.4.1.5 Integrated response options based on land management specifically for CDR**

3 While spreading of crushed minerals onto land as part of enhanced weathering can provide soil / plant
4 nutrients in nutrient-depleted soils, can increase soil organic carbon stocks and can help to replenish
5 eroded soil (Beerling et al. 2018), there is no literature on the global potential for addressing land
6 degradation.

7 Large-scale production of bioenergy can require significant amounts of land (Smith et al. 2016c;
8 Clarke and Jiang 2014b; Popp et al. 2017), much as 15 Mkm² in 2°C scenarios (Popp et al. 2017),
9 increasing pressures for land conversion and land degradation (Table 6.13). However, bioenergy
10 production can either increase (Robertson et al. 2017c; Mello et al. 2014a) or decrease (FAO 2011;
11 Lal 2014) soil organic matter, depending on where it is produced and how it is managed. These effects
12 are not included in the quantification in Table 6.41.

13 Table 6.41 summarises the impact on land degradation of options specifically for CDR, with
14 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
15 (not exhaustive) references upon which the evidence is based.

16 **Table 6.41 Effects on land degradation of response options specifically for CDR**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------------|--|-------------------|---|
| Enhanced weathering of minerals | Positive but not quantified | Low confidence | Beerling et al. 2018 |
| Bioenergy and BECCS | Negative impact on up to 15 Mkm ² | High confidence | Clarke et al. 2014a; Popp et al. 2017; Smith et al. 2016c |

17

18 **6.3.4.2 Integrated response options based on value chain management**

19 In this section, the impacts on land degradation of integrated response options based on value change
20 management are assessed.

21 **6.3.4.2.1 Integrated response options based on value chain management through demand management**

22
23 Dietary change and waste reduction both result in decreased cropland and pasture extent (Bajželj et al.
24 2014; Stehfest et al. 2009; Tilman and Clark 2014a), reducing the pressure for land degradation
25 (Table 6.15). Reduced post-harvest losses could spare 1.98 Mkm² of cropland globally (Kummu et al.
26 2012) meaning that land degradation pressure could be relieved from this land area through reduction
27 of post-harvest losses. The effects of material substitution on land degradation depend on
28 management practice; some forms of logging can lead to increased land degradation (Chapter 4).

1 Table 6.42 summarises the impact on land degradation of demand management options, with
 2 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
 3 (not exhaustive) references upon which the evidence is based.

4 **Table 6.42 Effects on land degradation of response options based on demand management**

| Integrated response option | Potential | Confidence | Citation |
|---|-----------------------|-------------------|--|
| Dietary change | 4-28 Mkm ² | High confidence | Alexander et al. 2016; Bajželj et al. 2014; Stehfest et al. 2009; Tilman and Clark 2014a |
| Reduced post-harvest losses | 1.98 Mkm ² | Medium confidence | Kummu et al. 2012a |
| Reduced food waste (consumer or retailer) | 7 Mkm ² | Medium confidence | Bajželj et al. 2014 |
| Material substitution | No global estimates | No evidence | |

5

6 **6.3.4.2.2 Integrated response options based on value chain management through supply** 7 **management**

8 There are no global estimates of the impact on land degradation of enhanced urban food systems,
 9 improved food processing, retailing, or improved energy use in food systems.

10 There is evidence that sustainable sourcing could reduce land degradation, as the explicit goal of
 11 sustainable certification programs is often to reduce deforestation or other unsustainable land uses.
 12 Over 4 Mkm² of forests are certified for sustainable harvesting (PEFC/FSC 2018), although it is not
 13 clear if all these lands would be at risk of degradation without certification. While the food price
 14 instability of 2007/2008 increased financial investment in crop expansion (especially through so-
 15 called land grabbing), and thus better management of supply chains might have reduced this amount,
 16 no quantification of the total amount of land acquired, nor the possible impact of this crop expansion
 17 on degradation, has been recorded (McMichael and Schneider 2011a; McMichael 2012).

18 Table 6.43 summarises the impact on land degradation of supply management options, with
 19 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
 20 (not exhaustive) references upon which the evidence is based.

21 **Table 6.43 Effects on land degradation of response options based on supply management**

| Integrated response option | Potential | Confidence | Citation |
|--|---------------------|-------------------|------------------|
| Sustainable sourcing | >4 Mkm ² | Low confidence | Auld et al. 2008 |
| Management of supply chains | No global estimates | No evidence | |
| Enhanced urban food systems | No global estimates | No evidence | |
| Improved food processing and retailing | No global estimates | No evidence | |
| Improved energy use in food systems | No global estimates | No evidence | |

22

23 **6.3.4.3 Integrated response options based on risk management**

24 In this section, the impacts on land degradation of integrated response options based on risk
 25 management are assessed.

26 Urban expansion has been identified as a major culprit in soil degradation in some countries; for
 27 example, urban expansion in China has now affected 0.2 Mkm², or almost one-sixth of the cultivated
 28 land total, causing an annual grain yield loss of up to 10 Mt, or around 5-6% of cropland production.
 29 Cropland production losses of 8-10% by 2030 are expected under model scenarios of urban expansion
 30 (Bren d'Amour et al. 2016). Pollution from urban development has included water and soil pollution
 31 from industry and wastes and sewage as well as acid deposition from increasing energy use in cities
 32 (Chen 2007a), all resulting in major losses to Nature's Contributions to People from urban conversion
 33 (Song and Deng 2015). Soil sealing from urban expansion is a major loss of soil productivity across

1 many areas. The World Bank has estimated that new city dwellers in developing countries will require
2 160–500 m² per capita, converted from non-urban to urban land (Barbero-Sierra et al. 2013a; Angel et
3 al 2005).

4 Degradation can be a driver leading to livelihood diversification (Batterbury 2001; Lestrelin and
5 Giordano 2007). Diversification has the potential to deliver some reversal of land degradation, if
6 diversification involves adding non-traditional crops or trees that may reduce the need for tillage
7 (Antwi-Agyei et al. 2014). China’s Sloping Land conversion programme has had livelihood
8 diversification benefits and is said to have prevented degradation of 93 thousand km² of land (Liu
9 et al. 2015). However, Warren (2002) provides conflicting evidence that more diverse-income
10 households had increased degradation on their lands in Niger, and Palacios et al. (2013) associate
11 landscape fragmentation with increased livelihood diversification in Mexico.

12 Use of local seeds may play a role in addressing land degradation due to the likelihood of local seeds
13 being less dependent on inputs such as chemical fertilisers or mechanical tillage; for example, in
14 India, local legumes are retained in seed networks while commercial crops like sorghum and rice
15 dominate food markets (Reisman 2017a). However, there are no global figures.

16 Disaster Risk Management systems can have some positive impacts on prevention and reversal of
17 land degradation, like the Global Drought Early Warning System (see section 6.3.3.3) (Pozzi et al.
18 2013).

19 Risk sharing instruments could have benefits for reduced degradation, but there are no global
20 estimates. Commercial crop insurance is likely to deliver no co-benefits for prevention and reversal of
21 degradation. One study found a 1% increase in farm receipts generated from subsidised farm
22 programmes (including crop insurance and others) increased soil erosion by 0.3 t ha⁻¹ (Goodwin and
23 Smith 2003). Wright and Wimberly (2013) found a 5310 km² decline in grasslands in the Upper
24 Midwest of the US during 2006-2010 due to crop conversion driven by higher prices and access to
25 insurance.

26 Table 6.44 summarises the impact on land degradation of risk management options, with confidence
27 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
28 exhaustive) references upon which the evidence is based.

29 **Table 6.44 Effects on land degradation of response options based on risk management**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------------|---|-------------------|---|
| Management of urban sprawl | >0.2 Mkm ² | Medium confidence | Chen 2007b; Zhang 2000 |
| Livelihood diversification | >0.1 Mkm ² | Low confidence | Liu and Lan 2015 |
| Use of local seeds | No global estimates | No evidence | |
| Disaster risk management | No global estimates | No evidence | Pozzi et al. 2013 |
| Risk sharing instruments | Variable, but negative impact on >5 thousand km ² in Upper Midwest USA | Low confidence | Goodwin and Smith 2003; Wright and Wimberly 2013 |

30

31 **6.3.5 Potential of the integrated response options for addressing food security**

32 In this section, the impacts of integrated response options on food security are assessed.

33 **6.3.5.1 Integrated response options based on land management**

34 In this section, the impacts on food security of integrated response options based on land management
35 are assessed.

1 **6.3.5.1.1 *Integrated response options based on land management in agriculture***

2 Increased food productivity has fed many millions of people, who could not have otherwise been fed.
3 Erisman et al. (2008) estimated that over 3 billion people worldwide could not have been fed without
4 increased food productivity arising from N fertilisation (Table 6.45).

5 Improved cropland management to achieve food security aims at closing yield gaps by increasing use
6 efficiency of essential inputs such as water and nutrients. Large production increases (45–70% for
7 most crops) are possible from closing yield gaps to 100% of attainable yield, by increasing fertiliser
8 use and irrigation, but overuse of nutrients could cause adverse environmental impacts (Mueller et al.
9 2012). This improvement can impact 1000 million people.

10 Improved grazing land management includes grasslands, rangelands and shrublands, and all sites on
11 which pastoralism is practiced. In general terms, continuous grazing may cause severe damage to
12 topsoil quality, through e.g. compaction. This damage may be reversed by short grazing exclusion
13 periods under rotational grazing systems (Greenwood and McKenzie 2001; Drewry 2006; Taboada et
14 al. 2011). Due to the widespread diffusion of pastoralism, improved grassland management may
15 potentially affect more than 1000 million people, many of them under subsistence agricultural
16 systems.

17 Meat, milk, eggs, and other animal products, including fish and other seafoods, will play an important
18 role in achieving food security (Reynolds et al. 2015). Improved livestock management with different
19 animal types and feeds may also impact one million people (Herrero et al. 2016). Ruminants are
20 efficient converters of grass into human edible energy and protein and grassland-based food
21 production can produce food with a comparable carbon footprint to mixed systems (O'Mara 2012b).
22 However, in the future, livestock production will increasingly be affected by competition for natural
23 resources, particularly land and water, competition between food and feed and by the need to operate
24 in a carbon-constrained economy (Thornton et al. 2009a).

25 Currently, over 1.3 billion people are on degrading agricultural land, and the combined impacts of
26 climate change and land degradation could reduce global food production by 10% by 2050. Since
27 agroforestry could help to address land degradation, up to 1.3 billion people could benefit in terms of
28 food security through agroforestry.

29 Agricultural diversification is not always economically viable; technological, biophysical,
30 educational, and cultural barriers may emerge that limit the adoption of more diverse farming systems
31 by farmers (Section 6.4.1). Nevertheless, diversification could benefit 1000 million people, many of
32 them under subsistence agricultural systems (Birthal et al. 2015; Massawe et al. 2016; Waha et al.
33 2018).

34 Cropland expansion during 1985 to 2005 was 17 thousand km² yr⁻¹ (Foley et al. 2005). Given that
35 cropland productivity (global average of 250 kg protein ha⁻¹ yr⁻¹ for wheat; (Clark and Tilman 2017)
36 is greater than that of grassland (global average of about 10 kg protein ha⁻¹ yr⁻¹ for beef/mutton; (Clark
37 and Tilman 2017), prevention of this conversion to cropland would have led to a loss of about 0.4 Mt
38 protein yr⁻¹ globally. Given an average protein consumption in developing countries of 25.5 kg protein
39 yr⁻¹ (equivalent to 70g person⁻¹ day⁻¹; FAO, 2018), this is equivalent to the protein consumption of
40 16.4 million people each year (Table 6.45).

41 Integrated water management provides direct benefits to food security by improving agricultural
42 productivity (Chapter 5; Tilman et al. 2011; Godfray and Garnett 2014), thereby potentially impacting
43 the livelihood and well-being of >1000 million people (Campbell et al. 2016) affected by hunger and
44 highly impacted by climate change. Increasing water availability and reliable supply of water for
45 agricultural production using different techniques of water harvesting, storage, and its judicious
46 utilisation through farm ponds, dams, and community tanks in rainfed agriculture areas have been
47 presented by Rao (2017) and Rivera-Ferre et al. (2016).

1 Table 6.45 summarises the impact on food security of options in agriculture, with confidence
 2 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 3 exhaustive) references upon which the evidence is based.

4 **Table 6.45 Effects on food security of response options in agriculture**

| Integrated response option | Potential | Confidence | Citation |
|--|--|-----------------|--|
| Increased food productivity | 3000 million people | High confidence | Erisman et al. 2008 |
| Improved cropland management | >1000 million people | Low confidence | Campbell et al. 2014; Lipper et al. 2014 |
| Improved grazing land management | >1000 million people | Low confidence | Herrero et al. 2016 |
| Improved livestock management | >1000 million people | Low confidence | Herrero et al. 2016 |
| Agroforestry | Up to 1300 million people | Low confidence | Sasha et al. 2018 |
| Agricultural diversification | >1000 million people | Low confidence | Birthal et al. 2015; Massawe et al. 2016; Waha et al. 2018 |
| Reduced grassland conversion to cropland | Negative impact on 16.4 million people | Low confidence | Clark and Tilman 2017; FAO, 2018 |
| Integrated water management | >1000 million people | High confidence | Campbell et al. 2016 |

5

6 **6.3.5.1.2 Integrated response options based on land management in forestry**

7 Forests play a major role in providing food to local communities (non-timber forest products,
 8 mushrooms, fodder, fruits, berries etc.), and diversify daily diets directly or indirectly through
 9 improving productivity, hunting, diversifying tree-cropland-livestock systems, and grazing in forests.
 10 Based on the extent of forest contributing to food supply, considering the people undernourished
 11 (Rowland et al. 2017; FAO, IFAD, and WFP, 2013), and the annual deforestation rate (Keenan et al.
 12 2015), the global potential to enhance food security is moderate for improved forest management and
 13 small for reduced deforestation (Table 6.46). The uncertainty of these global estimates is high. More
 14 robust qualitative and some quantitative estimates are available at regional level. For example,
 15 managed natural forests, shifting cultivation and agroforestry systems are demonstrated to be crucial
 16 to food security and nutrition for hundreds of million people in rural landscapes worldwide
 17 (Sunderland et al. 2013; Vira et al. 2015). According to Erb et al. (2016), deforestation would not be
 18 needed to feed the global population by 2050, in terms of quantity and quality of food. At local level,
 19 Cerri et al. (2018) suggested that reduced deforestation, along with integrated cropland-livestock
 20 management, would positively impact more than 120 million people in the Cerrado, Brazil. In Sub-
 21 Saharan Africa, where population and food demand are projected to continue to rise substantially,
 22 reduced deforestation may have strong positive effects on food security (Doelman et al. 2018).

23 Afforestation and reforestation negatively impact food security (Boysen et al. 2017b; Frank et al.
 24 2017; Kreidenweis et al. 2016b). It is estimated that large-scale afforestation plans could cause
 25 increases in food prices of 80% by 2050 (Kreidenweis et al. 2016b), and more general mitigation
 26 measures in the AFOLU sector can translate into a rise in undernourishment of 80–300 million people
 27 (Frank et al. 2017) (Table 6.16). For reforestation, the potential adverse side-effects with food security
 28 are smaller than afforestation, because forest regrows on recently deforested areas, and its impact
 29 would be felt mainly through impeding possible expansion of agricultural areas. On a smaller scale,
 30 forested land also offers benefits in terms of food supply, especially when forest is established on
 31 degraded land, mangroves, and other land that cannot be used for agriculture. For example, food from
 32 forests represents a safety-net during times of food and income insecurity (Wunder et al., 2014), and

1 wild harvested meat and freshwater fish provides 30-80% of protein intake from many rural
2 communities (McIntyre et al., 2016; Nasi et al., 2011).

3 Table 6.46 summarises the impact on food security of options in forestry, with confidence estimates
4 based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not exhaustive)
5 references upon which the evidence is based.

6 **Table 6.46 Effects on food security of response options in forestry**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|---|-------------------|--|
| Improved forest management | Positive impact on < 100 million people | Low confidence | FAO, IFAD & WFP, 2013; Rowland et al. 2017 |
| Reduced deforestation and degradation | Positive impact on < 1 million people | Low confidence | FAO, IFAD & WFP, 2013; Keenan et al. 2015; Rowland et al. 2017 |
| Reforestation and forest restoration | See afforestation | | |
| Afforestation | Negative impact on > 100 million people | Medium confidence | Boysen et al. 2017b; Frank et al. 2017; Kreidenweis et al. 2016b |

7

8 **6.3.5.1.3 Integrated response options based on land management of soils**

9 Increasing soil organic matter stocks can increase yield and improve yield stability (Lal 2006b; Pan et
10 al. 2009; Soussana et al. 2019), though this is not universally seen (Hijbeek et al., 2017). Lal (2006b)
11 concludes that crop yields can be increased by 20–70 kg ha⁻¹, 10–50 kg ha⁻¹ and 30–300 kg ha⁻¹ for
12 maize for wheat, rice and maize, respectively, for every 1 t C ha⁻¹ increase in soil organic carbon in
13 the root zone. Increasing soil organic carbon by 1 t C ha⁻¹ could increase food grain production in
14 developing countries by 32 Mt yr⁻¹ (Lal 2006b). Frank et al. (2017) estimate that soil carbon
15 sequestration could reduce calorie loss associated with agricultural mitigation measures by 65%,
16 saving 60–225 million people from undernourishment compared to a baseline without soil carbon
17 sequestration (Table 6.47).

18 Lal (1998) estimated the risks of global annual loss of food production due to accelerated erosion to
19 be as high as 190 Mt yr⁻¹ of cereals, 6 Mt yr⁻¹ of soybean, 3 Mt yr⁻¹ of pulses and 73 Mt yr⁻¹ of roots
20 and tubers. Considering only cereals, if we assume per-capita annual grain consumption in developing
21 countries to be 300 kg yr⁻¹ (estimated based on data included in Pradhan et al., 2013; FAO, 2018;
22 FAO et al., 2018; and World Bank 2018a), the loss of 190 Mt yr⁻¹ of cereals is equivalent to that
23 consumed by 633 million people, annually (Table 6.47).

24 Though there are biophysical barriers, such as access to appropriate water sources and limited
25 productivity of salt-tolerant crops, prevention / reversal of soil salinisation could benefit 1–100
26 million people (Qadir et al. 2013a). Soil compaction affects crop yields, so prevention of compaction
27 could benefit an estimated 1–100 million people globally (Anderson and Peters 2016).

28 Biochar on balance, could provide moderate benefits for food security by improving yields by 25% in
29 the tropics, but with more limited impacts in temperate regions (Jeffery et al. 2017), or through
30 improved water holding capacity and nutrient use efficiency (Chapter 5; Sohi 2012). These benefits
31 could, however, be tempered by additional pressure on land if large quantities of biomass are required
32 as feedstock for biochar production, thereby causing potential conflicts with food security (Smith
33 2016b). Smith (2016b) estimated that 0.4–2.6 Mkm² of land would be required for biomass feedstock
34 to deliver 2.57 GtCO₂e yr⁻¹ of CO₂ removal. If biomass production occupied 2.6 Mkm² of cropland,
35 equivalent to around 20% of the global cropland area, this could potentially have a large effect on
36 food security, although Woolf et al. (2010) argue that abandoned cropland could be used to supply
37 biomass for biochar, thus avoiding competition with food production. Similarly, Woods et al (2015)
38 estimate that 5-9 Mkm² of land is available for biomass production without compromising food

1 security and biodiversity, considering marginal and degraded land and land released by pasture
2 intensification (Table 6.47).

3 Table 6.47 summarises the impact on food security of soil-based options, with confidence estimates
4 based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not exhaustive)
5 references upon which the evidence is based.

6 **Table 6.47 Effects on food security of soil-based response options**

| Integrated response option | Potential | Confidence | Citation |
|---------------------------------------|--|-------------------|---|
| Increased soil organic carbon content | 60-225 million people | Low confidence | Frank et al. 2017 |
| Reduced soil erosion | 633 million people yr ⁻¹ | Low confidence | FAO, 2018; FAO et al. 2018; Lal 1998; Pradhan et al. 2013; World Bank 2018a |
| Reduced soil salinisation | 1-100 million people | Low confidence | Qadir et al. 2013a |
| Reduced soil compaction | 1-100 million people | Low confidence | Anderson and Peters 2016 |
| Biochar addition to soil | Range from positive impact in the tropics from biochar addition to soil to a maximum potential negative impact on >100 million people by worst-case conversion of 20% of global cropland | Low confidence | Jeffery et al. 2017; worse case negative impacts calculated from area values in Smith 2016b |

7

8 **6.3.5.1.4 Integrated response options based on land management across all/other ecosystems**

9 FAO (2015) calculated that damage from forest fires between 2003 and 2013 impacted a total of 49
10 thousand km² of crops with the vast majority in Latin America. Based on the world cereal yield in
11 2013 reported by Word Bank (2018b) (3.8 t ha⁻¹), the loss of 49 thousand km² of crops is equivalent to
12 18.6 Mt yr⁻¹ of cereals lost. Assuming annual grain consumption per capita to be 300 kg yr⁻¹
13 (estimated based on data included in Pradhan et al., 2013; FAO, 2018; FAO et al., 2018; and World
14 Bank 2018a), the loss of 18.6 Mt yr⁻¹ would remove cereal crops equivalent to that consumed by 62
15 million people (Table 6.48).

16 Landslides and other natural hazards affect 1–100 Million people globally, so preventing them could
17 provide food security benefits to this many people.

18 In terms of measures to tackle pollution, including acidification, Shindell et al. (2012) considered
19 about 400 emission control measures to reduce ozone and black carbon (BC). This strategy increases
20 annual crop yields by 30–135 Mt due to ozone reductions in 2030 and beyond. If annual grain
21 consumption per capita is assumed as 300 kg yr⁻¹ (estimated based on data included in Pradhan et al.,
22 2013; FAO, 2018; FAO et al., 2018; and World Bank 2018a), increase in annual crop yields by 30–
23 135 Mt feeds 100–450 million people.

24 There are no global data on the impacts of management of invasive species / encroachment on food
25 security.

26 Since large areas of converted coastal wetlands are used for food production (e.g., mangroves
27 converted for aquaculture; (Naylor et al. 2000b), restoration of coastal wetlands could displace food
28 production and damage local food supply, potentially leading to adverse impacts on food security,
29 though these effects are likely to be very small given that only 0.3% of human food comes from the

1 oceans and other aquatic ecosystems (Pimentel 2006), and that the impacts could be offset by careful
2 management, such as the careful siting of ponds within mangroves (Naylor et al. 2000b) (Table 6.46).

3 Around 14-20% (0.56–0.80 Mkm²) of the global 4 Mkm² of peatlands are used for agriculture, mostly
4 for meadows and pasture, meaning that if all of these peatlands were removed from production, 0.56–
5 0.80 Mkm² of agricultural land would be lost. Assuming livestock production on this land (since it is
6 mostly meadow and pasture) with a mean productivity of 9.8 kg protein ha⁻¹ yr⁻¹ (calculated from land
7 footprint of beef/mutton in (Clark and Tilman 2017)), and average protein consumption in developing
8 countries of 25.5 kg protein yr⁻¹ (equivalent to 70g person⁻¹ day⁻¹; FAO, 2018), this would be
9 equivalent to 21–31 million people no longer fed from this land (Table 6.46).

10 There are no global estimates on how biodiversity conservation improves nutrition (i.e. number of
11 nourished people). Biodiversity, and its management, is crucial for improving sustainable and
12 diversified diets (Global Panel on Agriculture and Food Systems for Nutrition 2016). Indirectly, the
13 loss of pollinators (due to combined causes, including the loss of habitats and flowering species)
14 would contribute to 1.42 million additional deaths per year from non-communicable and malnutrition-
15 related diseases, and 27.0 million lost disability-adjusted life-years (DALYs) per year (Smith et al.
16 2015). However, at the same time, some options to preserve biodiversity, like protected areas, may
17 potentially conflict with food production by local communities (Molotoks et al. 2017).

18 Table 6.48 summarises the impact on food security of response options in all/other ecosystems, with
19 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
20 (not exhaustive) references upon which the evidence is based.

21 **Table 6.48 Effects on food security of response options in all/other ecosystems**

| Integrated response option | Potential | Confidence | Citation |
|--|--|----------------|--|
| Fire management | ~62 million people | Low confidence | FAO 2015; FAO 2018; FAO et al. 2018; Pradhan et al. 2013; World Bank 2018a,b |
| Reduced landslides and natural hazards | 1-100 million people | Low confidence | Campbell 2015 |
| Reduced pollution including acidification | Increase annual crop yields 30-135 Mt globally; feeds 100-450 million people | Low confidence | Shindell et al. 2012; FAO, 2018; FAO et al., 2018; Pradhan et al. 2013; World Bank 2018a |
| Management of invasive species / encroachment | No global estimates | No evidence | |
| Restoration and reduced conversion of coastal wetlands | Very small negative impact but not quantified | Low confidence | |
| Restoration and reduced conversion of peatlands | Potential negative impact on 21-31 million people | Low confidence | Clark and Tilman 2017; FAO 2018 |
| Biodiversity conservation | No global estimates | No evidence | |

22

23 **6.3.5.1.5 Integrated response options based on land management specifically for CDR**

24 The spreading of crushed minerals on land as part of enhanced weathering on nutrient-depleted soils
25 can potentially increase crop yield by replenishing plant available silicon, potassium and other plant
26 nutrients (Beerling et al. 2018), but there are no estimates in the literature reporting the potential
27 magnitude of this effect on global food production.

28 Competition for land between bioenergy and food crops can lead to adverse side-effects for food
29 security. Many studies indicate that bioenergy could increase food prices (Calvin et al. 2014a; Popp et

1 al. 2017; Wise et al. 2009a). Only three studies were found linking bioenergy to the population at risk
2 of hunger; they estimate an increase in the population at risk of hunger of between 2 million and 150
3 million people (Table 6.49).

4 Table 6.49 summarises the impact on food security of response options specifically for CDR, with
5 confidence estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative
6 (not exhaustive) references upon which the evidence is based.

7 **Table 6.49 Effects on food security of response options specifically for CDR**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------------|---|-------------------|--|
| Enhanced weathering of minerals | No global estimates | No evidence | |
| Bioenergy and BECCS | Negative impact on up to 150 million people | Medium confidence | Baldos and Hertel 2014a; Fujimori et al. 2018b |

8 **6.3.5.2 Integrated response options based on value chain management**

9 In this section, the impacts on food security of integrated response options based on value change
10 management are assessed.

11 **6.3.5.2.1 Integrated response options based on value chain management through demand
12 management**

13 Dietary change can free up agricultural land for additional production (Bajželj et al. 2014; Stehfest et
14 al. 2009; Tilman and Clark 2014b) and reduce the risk of some diseases (Tilman and Clark 2014b;
15 Aleksandrowicz et al. 2016b) , with large positive impacts on food security (Table 6.50).

16 Kummu et al. (2012a) estimate that an additional billion people could be fed if food waste was halved
17 globally. This includes both post-harvest losses and retail and consumer waste, and measures such as
18 improved food transport and distribution could also contribute to this waste reduction (Table 6.50).

19 While no studies quantified the effect of material substitution on food security, the effects are
20 expected to be similar to reforestation and afforestation if the amount of material substitution leads to
21 an increase in forest area.

22 Table 6.50 summarises the impact on food security of demand management options, with confidence
23 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
24 exhaustive) references upon which the evidence is based.

25 **Table 6.50 Effects on food security of demand management options**

| Integrated response option | Potential | Confidence | Citation |
|---|-------------------------|-------------------|---|
| Dietary change | 821 million people | High confidence | Aleksandrowicz et al. 2016b; Tilman and Clark 2014b |
| Reduced post-harvest losses | 1000 million people | Medium confidence | Kummu et al. 2012 |
| Reduced food waste (consumer or retailer) | 700-1000 million people | Medium confidence | FAO 2018; Kummu et al. 2012 |
| Material substitution | No global estimates | No evidence | |

26

27 **6.3.5.2.2 Integrated response options based on value chain management through supply
28 management**

29 Since 810 million people are undernourished (FAO, 2018), this sets the maximum number of those
30 that could potentially benefit from sustainable sourcing or better management of supply chains.
31 Currently however, only 1 million people are estimated to benefit from sustainable sourcing (Tayleur

1 et al. 2017). For the latter, food price spikes affect food security and health; there are clearly
 2 documented effects of stunting among young children as a result of the 2007/2008 food supply crisis
 3 (de Brauw 2011; Arndt et al. 2012; Brinkman et al. 2010; Darnton-Hill and Cogill 2010) with a 10%
 4 increase in wasting attributed to the crisis in South Asia (Vellakkal et al. 2015). There is conflicting
 5 evidence on the impacts of different food price stability options for supply chains, and little
 6 quantification (Byerlee et al. 2006; del Ninno et al. 2007; Alderman 2010; Braun et al. 2014).
 7 Reduction in staple food prices due to price stabilisation resulted in more expenditure on other foods
 8 and increased nutrition (e.g., oils, animal products), leading to a 10% reduction in malnutrition among
 9 children in one study (Torlesse et al. 2003a). Comparison of two African countries shows that
 10 protectionist policies (food price controls) and safety nets to reduce price instability resulted in a 20%
 11 decrease in risk of malnutrition (Nandy et al. 2016). Models using policies for food aid and domestic
 12 food reserves to achieve food supply and price stability showed the most effectiveness of all options
 13 in achieving climate mitigation and food security goals (e.g. more effective than carbon taxes) as they
 14 did not exacerbate food insecurity and did not reduce ambitions for achieving temperature goals
 15 (Fujimori et al. 2018a).

16 For urban food systems, increased food production in cities combined with governance systems for
 17 distribution and access can improve food security, with a potential to produce 30% of food consumed
 18 in cities. The urban population in 2018 was 4.2 billion people, so 30% represents 1230 million people
 19 who could benefit in terms of food security from improved urban food systems (Table 6.51).

20 It is estimated that 500 million smallholder farmers depend on agricultural businesses in developing
 21 countries (World Bank, 2017), which sets the maximum number of people who could benefit from
 22 improved efficiency and sustainability of food processing, retail and agri-food industries.

23 Up to 2500 million people could benefit from increased energy efficiency in agriculture, based on the
 24 estimated number of people worldwide lacking access to clean energy and instead relying on biomass
 25 fuels for their household energy needs (IEA, 2014).

26 Table 6.51 summarises the impact on food security of supply management options, with confidence
 27 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 28 exhaustive) references upon which the evidence is based.

29 **Table 6.51 Effects on food security of supply management options**

| Integrated response option | Potential | Confidence | Citation |
|--|---------------------------|-------------------|---|
| Sustainable sourcing | > 1 million people | Low confidence | Tayleur et al. 2017 |
| Management of supply chains | > 1 million people | Low confidence | FAO 2018; Kummu et al. 2012 |
| Enhanced urban food systems | Up to 1260 million people | Low confidence | Benis and Ferrão 2017b; Padgham et al.; Specht et al. 2014; de Zeeuw & Drechsel 2015; |
| Improved food processing and retailing | 500 million people | Low confidence | World Bank 2017 |
| Improved energy use in food systems | Up to 2500 million people | Low confidence | IEA 2014 |

30

31 **6.3.5.3 Integrated response options based on risk management**

32 In this section, the impacts on food security of integrated response options based on risk management
 33 are assessed.

34 Evidence in the US indicates ambiguous trends between sprawl and food security; on one hand, most
 35 urban expansion in the US has primarily been on lands of low and moderate soil productivity with

1 only 6% of total urban land on highly productive soil. On the other hand, highly productive soils have
 2 experienced the highest rate of conversion of any soil type (Nizeyimana et al. 2001). Specific types of
 3 agriculture are often practiced in urban-influenced fringes, such as fruits, vegetables, and poultry and
 4 eggs in the US, the loss of which can have an impact on the types of nutritious foods available in
 5 urban areas (Francis et al. 2012b). China is also concerned with food security implications of urban
 6 sprawl, and a loss of 30 Mt of grain production from 1998–2003 in eastern China was attributed to
 7 urbanisation (Chen 2007b). However, overall global quantification has not been attempted.

8 Diversification is associated with increased welfare and incomes and decreased levels of poverty in
 9 several country studies (Arslan et al. 2018a; Asfaw et al. 2018). These are likely to have large food
 10 security benefits (Barrett et al. 2001; Niehof 2004), but there is little global quantification.

11 Local seed use can provide considerable benefits for food security because of the increased ability of
 12 farmers to revive and strengthen local food systems (McMichael and Schneider 2011b); studies have
 13 reported more diverse and healthy food in areas with strong food sovereignty networks (Coomes et al.
 14 2015b; Bisht et al. 2018). Women in particular may benefit from seed banks for low value but
 15 nutritious crops (Patnaik et al. 2017). Many hundreds of millions of smallholders still rely on local
 16 seeds and they provision many hundreds of millions of consumers (Altieri et al. 2012a; McGuire and
 17 Sperling 2016), so keeping their ability to do so through seed sovereignty is important. However,
 18 there may be lower food yields from local and unimproved seeds, so the overall impact of local seed
 19 use on food security is ambiguous (McGuire and Sperling 2016).

20 Disaster risk management approaches can have important impacts on reducing food insecurity, and
 21 current systems for drought warning and other storms currently reach over 100 million people. When
 22 these early warning systems can help farmers harvest crops in advance of impending weather events
 23 or otherwise make agricultural decisions to prepare for adverse events, there are likely to be positive
 24 impacts on food security (Fakhrudin et al. 2015). Surveys with farmers reporting food insecurity
 25 from climate impacts have indicated their strong interest in having such early warning systems
 26 (Shisanya and Mafongoya 2016). Additionally, famine early warning systems have been successful in
 27 Sahelian Africa to alert authorities of impending food shortages so that food acquisition and
 28 transportation from outside the region can begin, potentially helping millions of people (Genesio et al.
 29 2011; Hillbruner and Moloney 2012).

30 Risk sharing instruments are often aimed at sharing food supplies and reducing risk, and thus are
 31 likely to have important, but unquantified, benefits for food security. Crop insurance in particular has
 32 generally led to (modest) expansions in cultivated land area and increased food production (Claassen
 33 et al. 2011; Goodwin et al. 2004).

34 Table 6.52 summarises the impact on food security of risk management options, with confidence
 35 estimates based on the thresholds outlined in Table 6.53 in section 6.3.6, and indicative (not
 36 exhaustive) references upon which the evidence is based.

37 **Table 6.52 Effects on food security of risk management options**

| Integrated response option | Potential | Confidence | Citation |
|-----------------------------------|-------------------|-------------------|--|
| Management of urban sprawl | >1 million likely | Low confidence | Bren d'Amour et al. 2016; Chen 2017 |
| Livelihood diversification | >100 million | Low confidence | Morton 2007 |
| Use of local seeds | >100 million | Low confidence | Altieri et al. 2012a |
| Disaster risk management | > 100 million | Medium confidence | Genesio et al. 2011; Hillbruner and Moloney 2012 |
| Risk sharing instruments | >1 million likely | Low confidence | Claassen et al. 2011; Goodwin et al. 2004 |

1
2
3
4
5
6

6.3.6 Summarising the potential of the integrated response options across mitigation, adaptation, desertification land degradation and food security

Using the quantification provided in tables 6.13 to 6.52, the impacts are categorised as either positive or negative, and are designated as large, moderate and small according to the criteria given in Table 6.53⁶.

⁶ FOOTNOTE: Note that: 1) The response options often overlap, so are not additive. For example, increasing food productivity will involve changes to cropland, grazing land and livestock management, which in turn may include increasing soil carbon stocks. The response options cannot therefore be summed, nor regarded as entirely mutually exclusive interventions. 2) The efficacy of a response option for addressing the primary challenge for which it is implemented needs to be weighed against any co-benefits and adverse side-effects for the other challenges, e.g. if a response option has a major impact in addressing one challenge but results in relatively minor and manageable adverse-side effects for another challenge, it may remain a powerful response option despite the adverse side-effects, particularly if they can be minimised or managed. 3) Though the impacts of integrated response options have been quantified as far as possible in **Section 6.3**, there is no equivalence implied in terms benefits or adverse side-effects, either in number or in magnitude of the impact, i.e. one benefit *does not equal* one adverse side-effect. As a consequence: a) Large benefits for one challenge might outweigh relatively minor adverse side-effects in addressing another challenge, and b) Some response options may deliver mostly benefits with few adverse-side effects, but the benefits might be small in magnitude, i.e. the response options do no harm, but present only minor co-benefits. A number of benefits and adverse side-effects are context specific; the context specificity has been discussed in **section 6.2** and is further examined **Section 6.4.5.1**.

1 **Table 6.53 Key for criteria used to define magnitude of impact of each integrated response option**

| | Mitigation | Adaptation | Desertification | Land Degradation | Food |
|--------------------------|---|---|---|---|--|
| Large positive | More than 3 GtCO ₂ -eq yr ⁻¹ | Positively impacts more than around 25 million people | Positively impacts more than around 3 million km ² | Positively impacts more than around 3 million km ² | Positively impacts more than around 100 million people |
| Moderate positive | 0.3 to 3 GtCO ₂ -eq | 1 million to 25 million | 0.5 to 3 million km ² | 0.5 to 3 million km ² | 1 million to 100 million |
| Small positive | >0 | Under 1 million | >0 | >0 | Under 1 million |
| Negligible | 0 | No effect | No effect | No effect | No effect |
| Small negative | <0 | Under 1 million | <0 | <0 | Under 1 million |
| Moderate negative | -0.3 to -3 GtCO ₂ -eq | 1 million to 25 million | 0.5 to 3 million km ² | 0.5 to 3 million km ² | 1 million to 100 million |
| Large negative | More than -3 GtCO ₂ -eq yr ⁻¹ | Negatively impacts more than around 25 million people | Negatively impacts more than around 3 million km ² | Negatively impacts more than around 3 million km ² | Negatively impacts more than around 100 million people |

2 **Note:** All numbers are for global scale; all values are for technical potential. For mitigation, the target is set at
3 around the level of large single mitigation measure (about 1 GtC yr⁻¹ = 3.67 GtCO₂-eq yr⁻¹) (Pacala and Socolow
4 2004), with a combined target to meet 100 GtCO₂ in 2100, to go from baseline to 2°C (Clarke and Jiang 2014b). For
5 adaptation, numbers are set relative to the about 5 million lives lost per year attributable to climate change and a
6 carbon-based economy, with 0.4 million per year attributable directly to climate change. This amounts to 100 million
7 lives predicted to be lost between 2010 and 2030 due to climate change and a carbon-based economy (DARA 2012),
8 with the largest category representing 25% of this total. For desertification and land degradation, categories are set
9 relative to the 10-60 million km² of currently degraded land (Gibbs and Salmon 2015) with the largest category
10 representing 30% of the lower estimate. For food security, categories are set relative to the roughly 800 million
11 people currently undernourished (HLPE 2017) with the largest category representing around 12.5% of this total.
12

13 Tables 6.54 to 6.61 summarise the potentials of the integrated response options across mitigation,
14 adaptation, desertification, land degradation and food security. Cell colours correspond to the large,
15 moderate and small impact categories shown in Table 6.53.

16 As seen in tables 6.54 to 6.61, three response options across the 14 for which there are data for every
17 land challenge: *increased food productivity*, *agroforestry* and *increased soil organic carbon content*,
18 deliver large benefits across all five land challenges.

19 A further six response options: *improved cropland management*, *improved grazing land management*,
20 *improved livestock management*, *agroforestry*, *fire management* and *reduced post-harvest losses*,
21 deliver either large or moderate benefits for all land challenges.

22 Three additional response options: *dietary change*, *reduced food waste* and *reduced soil salinisation*,
23 each missing data to assess global potential for just one of the land challenges, deliver large or
24 moderate benefits to the four challenges for which there are global data.

25 Eight response options: *increased food productivity*, *reforestation and forest restoration*,
26 *afforestation*, *increased soil organic carbon content*, *enhanced mineral weathering*, *dietary change*,

1 *reduced post-harvest losses, and reduced food waste, have large mitigation potential (>3 GtCO₂e yr⁻¹)*
2 *without adverse impacts on other challenges.*

3 *Sixteen response options: increased food productivity, improved cropland management, agroforestry,*
4 *agricultural diversification, improved forest management, increased soil organic carbon content,*
5 *reduced landslides and natural hazards, restoration and reduced conversion of coastal wetlands,*
6 *reduced post-harvest losses, sustainable sourcing, management of supply chains, improved food*
7 *processing and retailing, improved energy use in food systems, livelihood diversification, use of local*
8 *seeds, and disaster risk management, have large adaptation potential at global scale (positively*
9 *affecting >25 million people) without adverse side-effects for other challenges.*

10 Thirty-three of the 40 response options can be applied without requiring land use change and limiting
11 available land. A large number of response options do not require dedicated land, including several
12 land management options, all value chain options, and all risk management options. Four options, in
13 particular, could greatly increase competition for land if applied at scale: *afforestation, reforestation,*
14 *and land used to provide feedstock for bioenergy (with or without BECCS) and biochar, with three*
15 *further options: reduced grassland conversion to croplands, restoration and reduced conversion of*
16 *peatlands and restoration and reduced conversion of coastal wetlands having smaller or variable*
17 *impacts on competition for land. Other options such as reduced deforestation and degradation,*
18 *restrict land conversion for other options and uses.*

19 Some response options can be more effective when applied together; for example, dietary change and
20 waste reduction expand the potential to apply other options by freeing as much as 25 Mkm² (4-25
21 Mkm² for dietary change; Alexander et al. 2016; Bajželj et al. 2014; Stehfest et al. 2009; Tilman and
22 Clark 2014b and 7 Mkm² for reduced food waste; Bajželj et al. 2014).

23 In terms of the categories of response options, most agricultural land management response options
24 (all except for reduced grassland conversion to cropland which potentially adversely affects food
25 security), deliver benefits across the five land challenges (Table 6.54). Among the forest land
26 management options, afforestation and reforestation have the potential to deliver large co-benefits
27 across all land challenges except for food security, where these options provide a threat due to
28 competition for land (Table 6.55). Among the soil-based response options, some global data are
29 missing, but none except biochar shows any potential for negative impacts, with that potential
30 negative impact arising from additional pressure on land if large quantities of biomass feedstock are
31 required for biochar production (Table 6.56). Where global data exists, most response options in
32 other/all ecosystems deliver benefits except for a potential moderate negative impact on food security
33 by restoring peatlands currently used for agriculture (Table 6.57). Of the two response options
34 specifically targeted at CDR, there are missing data for enhanced weathering of minerals for three of
35 the challenges, but large-scale bioenergy and BECCS shows a potential large benefit for mitigation,
36 but small to large adverse impacts on the other four land challenges (Table 6.58), mainly driven by
37 increased pressure on land due to feedstock demand.

38 While data allow the impact of material substitution to be assessed only for mitigation, the three other
39 demand-side response options: dietary change, reduced post-harvest losses and reduced food waste
40 provide large or moderate benefits across all challenges for which data exist (Table 6.59). For none of
41 the supply-side response options is data available to assess the impact on more than three of the land
42 challenges, but there are large to moderate benefits for all those for which data are available (Table
43 6.60). Data are not available to assess the impact of risk management-based response options on all of
44 the challenges, but there are small to large benefits for all of those for which data are available (Table
45 6.61).

1 **Table 6.54 Summary of direction and size of impact of land management options in agriculture on mitigation, adaptation, desertification, land degradation and**
2 **food security**

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | |
|----------------------------------|------------|------------|-----------------|------------------|---------------|--|
| Increased food productivity | | | | | | <p>Context and evidence base for magnitude of effect</p> <p>These estimates assume that increased food production is implemented sustainably (e.g. through sustainable intensification: Garnett et al. 2013b; Pretty et al. 2018) rather than through increasing external inputs, which can have a range of negative impacts. <u>Mitigation</u>: Large benefits (Table 6.13). <u>Adaptation</u>: Large benefits (Chapter 2; Table 6.21; Campbell et al. 2014). <u>Desertification</u>: Large benefits (Chapter 3; Table 6.29; Dai 2010). <u>Land degradation</u>: Large benefits (Chapter 4; Table 6.37; Clay et al., 1995). <u>Food security</u>: Large benefits (Chapter 5; Table 6.45; Godfray et al. 2010b; Tilman et al. 2011; Godfray and Garnett 2014).</p> |
| Improved cropland management | | | | | | <p><u>Mitigation</u>: Moderate benefits by reducing greenhouse gas emissions and creating soil carbon sinks (Chapter 2; Table 6.13; Smith et al. 2008, 2014a). <u>Adaptation</u>: Large benefits by improving the resilience of food crop production systems to future climate change (Chapter 2; Table 6.21; Porter et al. 2014). <u>Desertification</u>: Large benefits by improving sustainable use of land in dry areas (Chapter 3; Table 6.29; Bryan et al. 2009b; Chen et al. 2010). <u>Land degradation</u>: Large benefits by forming a major component of sustainable land management (Chapter 4; Table 6.37; Labrière et al. 2015). <u>Food security</u>: Large benefits by improving agricultural productivity for food production (Chapter 5; Table 6.45; Porter et al. 2014).</p> |
| Improved grazing land management | | | | | | <p><u>Mitigation</u>: Moderate benefits by increasing soil carbon sinks and reducing greenhouse gas emissions (Chapter 2; Table 6.13; Herrero et al. 2016). <u>Adaptation</u>: Moderate benefits by improving the resilience of grazing lands to future climate change (Chapter 2; Table 6.21; Porter et al. 2014). <u>Desertification</u>: Moderate benefits by tackling overgrazing in dry areas to reduce desertification (Chapter 3; Table 6.29; Archer et al. 2011). <u>Land degradation</u>: Large benefits by optimising stocking density to reduce land degradation (Chapter 4; Table 6.37; Table 6.45; Tighe et al. 2012). <u>Food security</u>: Large benefits by improving livestock sector productivity to increase food production (Chapter 5; Table 6.45; Herrero et al. 2016).</p> |
| Improved livestock management | | | | | | <p><u>Mitigation</u>: Moderate benefits by reducing greenhouse gas emissions, particularly from enteric methane and manure management (Chapter 2; Table 6.13; Smith et al. 2008, 2014a). <u>Adaptation</u>: Moderate benefits by improving resilience of livestock production systems to climate change (Chapter 2; Table 6.21; Porter et al. 2014). <u>Desertification</u>: Moderate benefits by tackling overgrazing in dry areas (Chapter 3; Table 6.29; Archer et al. 2011). <u>Land degradation</u>: Large benefits by reducing overstocking which can reduce land degradation (Chapter 4; Table 6.37; Table 6.45; Tighe et al. 2012). <u>Food security</u>: Large benefits by improving livestock sector productivity to increase food production (Chapter 5; Table 6.45; Herrero et al. 2016).</p> |
| Agroforestry | | | | | | <p><u>Mitigation</u>: Moderate benefits by increasing carbon sinks in vegetation and soils (Chapter 2; Table 6.13; Delgado 2010; Mbow et al. 2014a; Griscom et al. 2017a). <u>Adaptation</u>: Large benefits by improving the resilience of agricultural lands to climate change (Chapter 2; Table 6.21; Mbow et al. 2014a). <u>Desertification</u>: Large benefits through e.g. provides perennial vegetation in dry areas (Chapter 3; Table 6.29; Nair et al. 2010; Lal 2001a). <u>Land degradation</u>: Large benefits by stabilising soils through perennial</p> |

| | | | | | | |
|--|--|--------|--|--|--|---|
| | | | | | | vegetation (Chapter 4; Table 6.37; Narain et al. 1997; Lal 2001a). Food production: <i>Large benefits</i> since well-planned agroforestry can enhance productivity (Chapter 5; Table 6.45; Bustamante et al. 2014b; Sasha et al., 2018). |
| Agricultural diversification | | | | | | Agricultural diversification is a collection of practices aimed at deriving more crops or products per unit of area (e.g. intercropping) or unit of time (e.g. double cropping, ratoon crops etc.). Mitigation: <i>Limited benefits</i> (Table 6.13). Adaptation: <i>Large benefits</i> through improved household income (Pellegrini and Tasciotti 2014; Table 6.21). Desertification: <i>Moderate benefits</i> , limited by global dryland cropped area (Table 6.29). Land degradation: <i>Large benefits</i> by reducing pressure on land (Table 6.37; Lambin and Meyfroidt 2011). Food security: <i>Large benefits</i> for food security by provision of more diverse foods (Chapter 5; Table 6.45; BIRTHAL et al. 2015; Massawe et al. 2016; Waha et al. 2018). |
| Reduced grassland conversion to cropland | | N D | | | | Mitigation: <i>Moderate benefits</i> by retaining soil carbon stocks that might otherwise be lost. Historical losses of soil carbon have been on the order of 500 GtCO ₂ (Table 6.13; Sanderman et al. 2017). Mean annual global cropland conversion rates (1961–2003) have been 0.36% per year (Krause et al. 2017), i.e. around 47 thousand km ² yr ⁻¹ – so preventing conversion could potentially save moderate emissions of CO ₂ . Adaptation: No literature (Table 6.21). Desertification: <i>Limited benefits</i> by shifting from annual crops to permanent vegetation cover under grass in dry areas (Chapter 3; Table 6.29). Land degradation: <i>Limited benefits</i> by shifting from annual crops to permanent vegetation cover under grass (Chapter 4; Table 6.37). Food security: <i>Moderate negative impacts</i> , since more land is required to produce human food from livestock products on grassland than from crops on cropland, meaning that a shift to grassland could reduce total productivity and threaten food security (Chapter 5; Table 6.45; Clark and Tilman 2017). |
| Integrated water management | | | | | | Mitigation: <i>Moderate benefits</i> by reducing greenhouse gas emissions mainly in cropland and rice cultivation (Chapter 2; Table 6.13; Smith et al. 2008, 2014a). Adaptation: <i>Large benefits</i> by improving the resilience of food crop production systems to future climate change (Chapter 2; Table 6.21; Porter et al. 2014). Desertification: <i>Limited benefits</i> by improving sustainable use of land in dry areas (Chapter 3; Table 6.29). Land degradation: <i>Limited benefits</i> by forming a major component of sustainable land and water management (Chapter 4; Table 6.37). Food security: <i>Large benefits</i> by improving agricultural productivity for food production (Chapter 5; Table 6.45; Tilman et al. 2011; Godfray and Garnett 2014). |

1 **Note:** Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
 2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.
 3

4 **Table 6.55 Summary of direction and size of impact of land management options in forests on mitigation, adaptation, desertification, land degradation and food**
 5 **security**

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | Context and evidence base for magnitude of effect |
|----------------------------|------------|------------|-----------------|------------------|---------------|---|
| Improved forest | | | | | | Mitigation: <i>Moderate benefits</i> by conserving and enhancing carbon stocks in forests and long-lived products, through for example, selective logging (Table 6.14; Smith et al. 2014a). Adaptation: <i>Large benefits</i> , including through improving ecosystem |

| | | | | | | |
|---------------------------------------|------------|------------|-----------|-----------|------------|--|
| management | Light Blue | Dark Blue | Dark Blue | Dark Blue | Light Blue | functionality and services, with mostly qualitative evidence at global scale and more robust estimates at regional level and local scale (Table 6.22; Locatelli et al. 2015d). <u>Desertification and land degradation</u> : Large benefits by helping to stabilise land and regulate water and microclimate (Chapters 3 and 4; Tables 6.30 and 6.38; Locatelli et al. 2015d). <u>Food security</u> : Moderate benefits with mostly qualitative estimate at global level, by providing food to local communities, and diversify daily diets (Chapter 5; Table 6.46). |
| Reduced deforestation and degradation | Dark Blue | Light Blue | Dark Blue | Dark Blue | Light Blue | <u>Mitigation</u> : Large benefits by maintaining carbon stocks in forest ecosystems (Chapter 2; Table 6.14). <u>Adaptation</u> Moderate benefits at global scale when effect is cumulated till the end of the century; local scale, co-benefits between REDD+ and adaptation of local communities can be more substantial (Long 2013; Morita & Matsumoto 2017), even if often difficult to quantify and not explicitly acknowledged (McElwee et al. 2017a; Table 6.22). <u>Desertification and land degradation</u> : Large benefits at global scale when effects are cumulated for at least 20 years, e.g. through reduced soil erosion (Borrelli et al. 2017; Tables 6.30 and 6.38). The uncertainty of these global estimates is high, while more robust qualitative and some quantitative estimates are available at regional level. <u>Food security</u> : Small benefits ; difficult to quantify at global level (Chapter 5; Table 6.46). |
| Reforestation and forest restoration | Dark Blue | Dark Blue | Dark Blue | Dark Blue | Orange | <u>Mitigation</u> : Large benefits by rebuilding the carbon stocks in forest ecosystems, although decreases in surface albedo can reduce the net climate benefits, particularly in areas affected by seasonal snow cover (Chapter 2; Table 6.14; Sonntag et al. 2016; Mahmood et al. 2014). <u>Adaptation</u> : Large benefits by provision of Nature’s Contributions to People, including improving ecosystem functionality and services, providing microclimatic regulation for people and crops, wood and fodder as safety nets, soil erosion protection and soil fertility enhancement for agricultural resilience, coastal area protection, water and flood regulation (Locatelli et al. 2015d; Table 6.22). <u>Desertification</u> : Large benefits through restoring forest ecosystems in dryland areas (Chapter 3; Table 6.30; Idris Medugu et al. 2010a; Salvati et al. 2014b). <u>Land degradation</u> : Large benefits by re-establishment of perennial vegetation (Chapter 4; Table 6.38; Ellison et al. 2017b). <u>Food security</u> : Moderate negative impacts due to potential competition for land for food production (Chapter 5; Table 6.46; Frank et al. 2017). |
| Afforestation | Dark Blue | Dark Blue | Dark Blue | Dark Blue | Dark Red | <u>Mitigation</u> : Large benefits for mitigation (Chapter 2; Table 6.14), especially if it occurs in the tropics and in areas that are not significantly affected by seasonal snow cover. <u>Adaptation</u> : Large benefits on adaptation (Chapter 2; Table 6.22; Kongsager et al. 2016; Reyer et al. 2009). <u>Desertification</u> : Large benefits by providing perennial vegetation in dry areas to help control desertification (Chapter 3; Table 6.30; Idris Medugu et al. 2010a; Salvati et al. 2014b). <u>Land degradation</u> : Large benefits by stabilising soils through perennial vegetation (Chapter 4; Table 6.38; Lal 2001a). <u>Food security</u> : Large negative impacts due to competition for land for food production (Chapter 5; Table 6.46; Kreidenweis et al. 2016b; Smith et al. 2013b). |

1 Note: Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
 2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.

3 **Table 6.56 Summary of direction and size of impact of soil-based land management options on mitigation, adaptation, desertification, land degradation and food**
 4 **security**

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | |
|---------------------------------------|------------|------------|-----------------|------------------|---------------|---|
| | | | | | | Context and evidence base for magnitude of effect |
| Increased soil organic carbon content | | | | | | <u>Mitigation</u> : Large benefits by creating soil carbon sinks (Table 6.15). <u>Adaptation</u> : Large benefits by improving resilience of food crop production systems to climate change (Chapter 2; Table 6.24; IPBES 2018). <u>Desertification</u> : Large benefits by improving soil health and sustainable use of land in dry areas (Chapter 3; Table 6.31; D’Odorico et al. 2013). <u>Land degradation</u> : Large benefits since it forms a major component of recommended practices for sustainable land management (Chapter 4; Table 6.39; Altieri and Nicholls 2017). <u>Food security</u> : Large benefits since it can increase yield and yield stability to enhance food production, though this is not always the case (Chapter 5; Table 6.47; Pan et al. 2009; Soussana et al. 2019; Hijbeek et al., 2017; Schjøning et al., 2018). |
| Reduced soil erosion | | | | | | <u>Mitigation</u> : Large benefits or large negative impacts , since the final fate of eroded material is still debated, at the global level it is debated whether it is a large source or a large sink (Chapter 2; Table 6.15; Hoffmann et al. 2013). <u>Adaptation</u> : Large benefits since soil erosion control prevents <u>desertification</u> (large benefits) and <u>land degradation</u> (large benefits), thereby improving the resilience of agriculture to climate change (Chapter 2, 3 and 4; Table 6.23, 6.30 and 6.39; Lal 1998; FAO and ITPS 2015). <u>Food security</u> : Large benefits mainly through the preservation of crop productivity (Chapter 5; Table 6.47; Lal 1998). |
| Reduced soil salinisation | N D | | | | | Techniques to prevent and reverse soil salinisation include groundwater management by drainage systems and/or crop rotation and use of amendments to alleviate soil sodicity. <u>Mitigation</u> : There are no studies to quantify the global impacts (Table 6.15). <u>Adaptation</u> : Moderate benefits by allowing existing crop systems to be maintained, reducing the need to abandon land (Table 6.23; UNCTAD 2011; Dagar et al. 2016b). <u>Desertification</u> and <u>land degradation</u> : Moderate benefits since soil salinisation is a main driver of both desertification and land degradation (Chapters 3 and 4; Tables 6.31 and 6.39; Rengasamy 2006; Dagar et al. 2016b). <u>Food security</u> : Moderate benefits by maintaining existing cropping systems and helping to close yield gaps in rainfed crops (Table 6.47). |
| Reduced soil compaction | N D | | N D | | | Techniques to prevent and reverse soil compaction are based on the combination of suitable crop rotations, tillage and regulation of agricultural traffic (Hamza and Anderson 2005b). <u>Mitigation</u> : The global mitigation potential has not been quantified (Table 6.15; Chamen et al. 2015a; Epron et al. 2016; Tullberg et al. 2018b). <u>Adaptation</u> : Limited benefits by improving productivity but on relatively small global areas (Table 6.22). <u>Desertification</u> : no global data (Table 6.31). <u>Land degradation</u> : Large benefits since soil compaction is a main driver of land degradation (Table 6.39; FAO and ITPS 2015). <u>Food security</u> : Moderate benefits by helping to close yield gaps where compaction is a limiting factor (Table 6.47; Anderson and Peters 2016). |
| Biochar addition to soil | | N D | N D | | | <u>Mitigation</u> : Large benefits by increasing recalcitrant carbon stocks in the soil (Chapter 2; Table 6.15; Smith 2016b; Fuss et al. 2018b; IPCC 2018). <u>Adaptation</u> : There are no global estimates of the impact of biochar on climate adaptation (Table 6.23). <u>Desertification</u> : There are no global estimates of the impact of biochar on desertification (Table 6.31). <u>Land degradation</u> : Limited benefits by improving the soil water holding capacity, nutrient use efficiency, and potentially ameliorating heavy metal pollution (Table 6.39; Sohi 2012). <u>Food security</u> : Limited benefits by increasing crop yields in the tropics (though not in temperate regions; Jeffery et al. 2017), but potentially Large negative impacts by creating additional pressure on land if large quantities of biomass |

feedstock are required for biochar production (Table 6.47).

1 Note: Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.

3 **Table 6.57 Summary of direction and size of impact of land management in all/other ecosystems on mitigation, adaptation, desertification, land degradation and**
4 **food security**

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | Context and evidence base for magnitude of effect |
|---|------------|------------|-----------------|------------------|---------------|--|
| Fire management | Dark blue | Light blue | Dark blue | Dark blue | Light blue | <u>Mitigation:</u> Large benefits by reduced size, severity, and frequency of wildfires, thereby preventing emissions and preserving carbon stocks (Table 6.16; Chapter 2, Cross-Chapter Box 3; Arora and Melton 2018). <u>Adaptation:</u> Moderate benefits by reducing mortality attributable to landscape fire smoke exposure, fire management provides adaptation benefits (Table 6.24; Doerr and Santín 2016; Johnston et al. 2012; Shannon et al., 2016). <u>Desertification:</u> Large benefits since control of wildfires and long-term maintenance of tree stock density protects against soil erosion (Table 6.32; Neary et al. 2009a; Arora and Melton 2018). <u>Land degradation:</u> Large benefits by stabilising forest ecosystems (Table 6.40; Neary et al. 2009a; Arora and Melton 2018). <u>Food security:</u> Moderate benefits by maintaining forest food product availability and preventing fire expansion to agricultural land (Table 6.48; FAO 2015; FAO 2018; FAO et al., 2018; Pradhan et al., 2013; World Bank 2018a,b). |
| Reduced landslides and natural hazards | Light blue | Dark blue | Light blue | Dark blue | Light blue | <u>Mitigation:</u> The prevention of landslides and natural hazards benefits mitigation, but because of the limited impact on GHG emissions and eventual preservation of topsoil carbon stores, the impact is estimated to be small globally (Table 6.16; IPCC AR5 WG2, Chapter 14). <u>Adaptation:</u> Provides structural/physical adaptations to climate change (Table 6.24; IPCC AR5 WG2, Chapter 14). <u>Desertification:</u> Due to the small global areas affected within global drylands, the benefits for desertification control are limited (Chapter 3; Table 6.32). <u>Land degradation:</u> Since landslides and natural hazards are among the most severe degradation processes, prevention will have a large positive impact on land degradation (Chapter 4; Table 6.40; FAO and ITPS 2015). <u>Food security:</u> In countries in which mountain slopes are cropped for food, such as in the Pacific Islands (Campbell 2015), the management and prevention of landslides can deliver benefits for food security, though the global areas are limited (Table 6.48). |
| Reduced pollution including acidification | Green | Light blue | Light blue | Light blue | Dark blue | <u>Mitigation:</u> Large benefits since measures to reduce emissions of Short-Lived Climate Pollutants (SLCPs) can slow projected global mean warming (UNEP and WMO 2011), with early intervention providing 0.5°C cooling by 2050 (Table 6.16; UNEP and WMO 2011). But moderate negative impacts are also possible since reduced reactive N deposition could decrease terrestrial carbon uptake (Table 6.16). <u>Adaptation:</u> Moderate benefits since controlling PM2.5 and ozone improves human health (Table 6.24; Anenberg et al. 2012). <u>Desertification:</u> Moderate benefits since salinisation, pollution, and acidification are stressors for desertification (Table 6.32; Oldeman et al. 1991). <u>Land degradation:</u> Moderate benefits since acid deposition is a significant driver of land degradation (Table 6.40; Oldeman et al. 1991; Smith et al. 2015). <u>Food security:</u> Large benefits since ozone is harmful to crops, so measures to reduce air pollution would be expected to increase crop production (Table 6.48; Shindell et al. 2012; Pradhan |

| | | | | | | |
|--|--------|--------|--------|--------|--------|---|
| | | | | | | et al., 2013; FAO, 2018; FAO et al., 2018; World Bank 2018a). |
| Management of invasive species / encroachment | N D | N D | N D | N D | N D | There is no literature that assesses the global potential of management of invasive species on <u>mitigation, adaptation, desertification, land degradation</u> or on <u>food security</u> (Table 6.16; Table 6.24; Table 6.33; Table 6.40; Table 6.48). |
| Restoration and reduced conversion of coastal wetlands | | | | | | <u>Mitigation</u> : Large benefits since coastal wetland restoration and avoided coastal wetland impacts deliver moderate carbon sinks by 2030 (Table 6.16; Griscom et al. 2017a). <u>Adaptation</u> : Large benefits by providing a natural defence against coastal flooding and storm surges by dissipating wave energy, reducing erosion and by helping to stabilise shore sediments (Table 6.24). <u>Desertification</u> : There is likely negligible impact of coastal wetland restoration for prevention of desertification (Table 6.32). <u>Land degradation</u> : Limited benefits since large areas of global coastal wetlands are degraded (Lotze et al. 2006; Griscom et al. 2017a; Table 6.40). <u>Food security</u> : Small benefits to small adverse impacts since large areas of converted coastal wetlands are used for food production (e.g. mangroves converted for aquaculture), restoration could displace food production and damage local food supply, though mangrove restoration can also restore local fisheries (Table 6.48; Naylor et al. 2000b). |
| Restoration and reduced conversion of peatlands | | N D | | | | <u>Mitigation</u> : Moderate benefits since avoided peat impacts and peat restoration deliver moderate carbon sinks by 2030 (Table 6.16; Griscom et al. 2017a), though there can be increases in methane emissions after restoration (Jauhiainen et al. 2008). <u>Adaptation</u> : Likely to be benefits by regulating water flow and preventing downstream flooding (Table 6.24; Munang et al. 2014a), but the global potential has not been quantified. <u>Desertification</u> : No impact since peatlands occur in wet areas and deserts in dry areas. <u>Land degradation</u> : Moderate benefits since large areas of global peatlands are degraded (Table 6.40; Griscom et al. 2017a). <u>Food security</u> : Moderate adverse impacts since restoration of large areas of tropical peatlands and some northern peatlands that have been drained and cleared for food production, could displace food production and damage local food supply (Table 6.48). |
| Biodiversity conservation | | | N D | N D | N D | <u>Mitigation</u> : Moderate benefits from carbon sequestration in protected areas (Table 6.16; Calvin et al. 2014a). <u>Adaptation</u> : Moderate benefits – likely many millions benefit adaptation and resilience of local communities to climate change (Table 6.24; CBD, 2008), though global potential is poorly quantified. <u>Desertification</u> : No global data (Table 6.32). <u>Land degradation</u> : No global data (Table 6.40). <u>Food security</u> : No global data (Table 6.48). |

1 Note: Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.

3 **Table 6.58 Summary of direction and size of impact of land management options specifically for CDR on mitigation, adaptation, desertification, land degradation
4 and food security**

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | Context and evidence base for magnitude of effect |
|----------------------------|------------|------------|-----------------|------------------|---------------|---|
| Enhanced | | N | N | | N | <u>Mitigation</u> : Moderate to large benefits by removing atmospheric CO ₂ (Table 6.17; Lenton 2010; Smith et al. 2016b; Taylor et al. |

| | | | | | | |
|------------------------|--|---|---|--|---|---|
| weathering of minerals | | D | D | | D | 2016b). Adaptation: There is no literature to assess the global impacts of enhanced mineral weathering on adaptation (Table 6.25) nor on desertification (Table 6.33). Land degradation: <i>Limited benefits</i> expected since ground minerals can increase pH where acidification is the driver of degradation (Table 6.41; Taylor et al. 2016b). Food security: Though there may be co-benefits for food production (Beerling et al. 2018), these have not been quantified globally (Table 6.49). |
| Bioenergy and BECCS | | | | | | Mitigation: <i>Large benefits</i> of large-scale bioenergy and BECCS by potential to remove large quantities of CO ₂ from the atmosphere (Table 6.17). Adaptation: <i>Limited adverse impacts</i> of large-scale bioenergy and BECCS by increasing pressure on land (Table 6.25). Desertification: Moderate adverse impacts of large-scale bioenergy and BECCS through increased pressure on land (Table 6.33). Land degradation: <i>Large adverse impacts</i> of large-scale bioenergy and BECCS through increased pressure on land (Table 6.41). Food security: <i>Large adverse impacts</i> of large-scale bioenergy and BECCS through increased competition for land for food (Table 6.49). These potentials and effects assume large areas of bioenergy crops resulting in large mitigation potentials (i.e. >3 GtCO ₂ yr ⁻¹). The sign and magnitude of the effects of bioenergy and BECCS depends on the scale of deployment, the type of bioenergy feedstock, which other response options are included, and where bioenergy is grown (including prior land use and indirect land use change emissions). For example, limiting bioenergy production to marginal lands or abandoned cropland would have negligible effects on biodiversity, food security, and potentially small co-benefits for land degradation; however, the benefits for mitigation would also be smaller (Cross-Chapter Box 7 on Bioenergy (Chapter 6); Table 6.13). |

1 Note: Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
 2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.
 3

4 **Table 6.59 Summary of direction and size of impact of demand management options on mitigation, adaptation, desertification, land degradation and food security**

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | Context and evidence base for magnitude of effect |
|-----------------------------|------------|------------|-----------------|------------------|---------------|---|
| Dietary change | | N D | | | | Mitigation: <i>Large benefits</i> for mitigation by greatly reducing GHG emissions (Chapter 5; Table 6.18). Adaptation: While it would be expected to help with adaptation by reducing agricultural land area, there are no studies providing global quantifications (Table 6.26). Desertification: Potential <i>moderate benefits</i> by decreasing pressure on land (restricted by relatively limited global area; Table 6.34). Land degradation: <i>Large benefits</i> by decreasing pressure on land (Table 6.42). Food security: Large benefits by decreasing competition for land allowing more food to be produced from less land (Table 6.50). |
| Reduced post-harvest losses | | | | | | Mitigation: <i>Large benefits</i> by reducing food sector GHG emissions and reducing area required to produce the same quantity of food (Table 6.18), though increased use of refrigeration could increase emissions from energy use. Adaptation: <i>Large benefits</i> by reducing pressure on land (Table 6.26). Desertification and land degradation: <i>Moderate benefits</i> for both by reducing pressure on land (Table 6.34; Table 6.42). Food security: <i>Large benefits</i> since most of the food wasted in developing countries arises from post-harvest losses (Chapter 5; Table 6.50; Ritzema et al. 2017). |

| | | | | | | |
|---|--|--------|--------|--------|--------|--|
| Reduced food waste (consumer or retailer) | | N D | | | | Mitigation: <i>Large benefits</i> by reducing food sector GHG emissions and reducing area required to produce the same quantity of food (Table 6.18). Adaptation: While it would be expected to help with adaptation by reducing agricultural land area, there are no studies quantifying global adaptation impacts (Table 6.26). Desertification: <i>Moderate benefits</i> by reducing pressure on land (Table 6.34). Land degradation: <i>Large benefits</i> by reducing pressure on land (Table 6.42). Food security: <i>Large benefits</i> since 30% of all food produced globally is wasted (Table 6.50; Kummu et al. 2012). |
| Material substitution | | N D | N D | N D | N D | Mitigation: <i>Moderate benefits</i> through long-lived carbon storage, and by substitution of materials with higher embedded GHG emissions (Table 6.18). No global studies available to assess the quantitative impact on adaptation, desertification, land degradation or food security (Table 6.26; Table 6.34; Table 6.42; Table 6.50). |

1 **Note:** Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.
3

3

4 **Table 6.60 Summary of direction and size of impact of supply management options on mitigation, adaptation, desertification, land degradation and food security**

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | Context and evidence base for magnitude of effect |
|-----------------------------|------------|------------|-----------------|------------------|---------------|--|
| Sustainable sourcing | N D | | N D | | | Mitigation: No studies available to assess the global impact (Table 6.19). Adaptation: <i>Moderate benefits</i> by diversifying and increasing flexibility in the food system to climate stressors and shocks while simultaneously creating economic alternatives for the poor (thereby strengthening adaptive capacity) and lowering expenditures of food processors and retailers by reducing losses (Chapter 5; Table 6.27; Muller et al. 2017a). Desertification: No studies available to assess the global impact (Table 6.35; Table 6.43). Land degradation: Potentially <i>large benefits</i> , as over 4 Mkm ² currently certified for sustainable forest production, which could increase in future (Table 6.44). Food security: <i>Moderate benefits</i> by diversifying markets and developing value-added products in the food supply system, by increasing its economic performance and revenues to local farmers (Reidsma et al. 2010), by strengthening the capacity of food production chains to adapt to future markets and to improve income of smallholder farmers (Chapter 5; Table 6.51; Murthy and Madhava Naidu 2012). It may also provide more direct links between producers and consumers. |
| Management of supply chains | N D | | N D | N D | | Mitigation: There are no studies assessing the mitigation potential globally (Table 6.19). Adaptation: <i>Large benefits</i> by improving resilience to price increases or reducing volatility of production (Table 6.27; Fafchamps et al. 1998; Haggblade et al. 2017). Desertification and land degradation: No studies assessing global potential (Table 6.35; Table 6.43). Food security: <i>Moderate benefits</i> through helping to manage food price increases and volatility (Table 6.51; Vellakkal et al. 2015; Arndt et al. 2016). |
| Enhanced urban food | N D | N D | N D | N D | | There are no studies that assess the global potential to contribute to mitigation, adaptation, desertification or land degradation (Table 6.19; Table 6.27; Table 6.35; Table 6.43). Food security: <i>Large benefits</i> by increasing food access to urban dwellers and |

| | | | | | | |
|--|--|--|--------|--------|--|---|
| systems | | | | | | shortening of supply chains (Chapter 5; Table 6.51; Chappell et al. 2016). |
| Improved food processing and retailing | | | N D | N D | | <u>Mitigation</u> : Moderate benefits through reduced energy consumption, climate-friendly foods and reduced GHG emissions from transportation (Avetisyan et al. 2014), waste (Porter et al. 2016b), and energy use (Table 6.19; Mohammadi et al. 2014; Song et al. 2017). <u>Adaptation</u> : Large benefits among poor farmers through reduced costs and improved resilience (Table 6.27). <u>Desertification</u> and <u>land degradation</u> : There are no studies assessing global potential (Table 6.35; Table 6.43). <u>Food security</u> : Large benefits by supporting healthier diets and reducing food loss and waste (Chapter 5; Table 6.51; Garnett 2011). |
| Improved energy use in food systems | | | N D | N D | | <u>Mitigation</u> : Moderate benefits by reducing GHG emissions through decreasing use of fossil fuels and energy-intensive products, though the emission reduction is not accounted for in the AFOLU sector (Table 6.19; Smith et al. 2014a; IPCC AR5 WG3 Chapter 11). <u>Adaptation</u> : Large benefits for small farmers by reducing costs and increasing their resilience to climate change (Table 6.27). <u>Desertification</u> and <u>land degradation</u> : There are no studies assessing global potential (Table 6.35; Table 6.43). <u>Food security</u> : Large benefits , largely by improving efficiency for 2.5 million people still using traditional biomass for energy (Chapter 5; Table 6.51). |

1 Note: Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
 2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.
 3
 4

Table 6.61 Summary of direction and size of impact of risk management options on mitigation, adaptation, desertification, land degradation and food security

| Integrated response option | Mitigation | Adaptation | Desertification | Land degradation | Food security | Context and evidence base for magnitude of effect |
|----------------------------|------------|------------|-----------------|------------------|---------------|---|
| Management of urban sprawl | N D | | | | | <u>Mitigation</u> : There are no studies assessing the global potential (Table 6.20). <u>Adaptation</u> : Moderate benefits - though poorly quantified globally, likely to affect many millions of people (Table 6.28). <u>Desertification</u> : Limited benefits - though poorly quantified globally, 5000 km ² is at risk from urban sprawl in Spain alone; (Table 6.36). <u>Land degradation</u> : Limited benefits - though poorly quantified globally, urban sprawl effects millions of ha of land (Table 6.44). <u>Food security</u> : Moderate benefits estimated from impacts on food supply in models (Table 6.52; Bren d'Amour et al. 2016). |
| Livelihood diversification | N D | | | | | <u>Mitigation</u> : There are no studies assessing the global potential (Table 6.20). <u>Adaptation</u> : Large benefits through helping households to buffer income fluctuations and providing a broader range of options for the future (Table 6.28; Ahmed and Stepp 2016b; Thornton and Herrero 2014). <u>Desertification</u> : There are no studies assessing the global potential, although there are anecdotal reports of limited benefits from improved land management resulting from diversification (Batterbury 2001; Herrmann and Hutchinson 2005; Stringer et al. 2009) (Table 6.36). <u>Land degradation</u> : Limited benefits , for example through improved land use mosaics (Ribeiro et al 2013), larger-scale adoption in China's Sloping Land Conversion program to diversify income and reduce degradation has impacted 0.1 Mkm ² (Liu and Lan 2015; Table 6.44). <u>Food security</u> : Large benefits since many of the world's 700 million smallholders practice diversification, helping to provide economic access to food (Table 6.52; Morton 2007). |

| | | | | | | |
|--------------------------|--------|--|--------|--------|--|--|
| Use of local seeds | N D | | N D | N D | | <p><u>Mitigation</u>: There are no studies assessing the global potential (Table 6.19). <u>Adaptation</u>: Large benefits given that 60 to 100% of seeds used in various countries of the global South are likely local farmer-bred (non-commercial) seed and moving to the use of commercial seed would increase costs considerably for these farmers. Seed networks and banks protect local agrobiodiversity and landraces, which are important to facilitate adaptation, and can provide crucial lifelines when crop harvests fail (Table 6.28; Louwaars 2002; Howard 2015; Coomes et al. 2015b; van Niekerk and Wynberg 2017b; Vasconcelos et al. 2013; Reisman 2017). <u>Desertification</u> and <u>land degradation</u>: There are no studies assessing global potential (Table 6.36; Table 6.44). <u>Food security</u>: Large benefits since local seeds increases the ability of farmers to revive and strengthen local food systems; several studies have reported more diverse and healthy food in areas with strong food sovereignty networks (Table 6.52; Coomes et al. 2015b; Bisht et al. 2018).</p> |
| Disaster risk management | N D | | N D | N D | | <p><u>Mitigation</u>: There are no studies to assess the global mitigation potential of different DRM approaches (Table 6.19). <u>Adaptation</u>: Large benefits due to widespread use of Early Warning Systems that reach hundreds of millions (Table 6.28; Hillbruner and Moloney 2012; Mahmud and Prowse 2012; Birkmann et al. 2015b). <u>Desertification</u> and <u>land degradation</u>. There are no studies assessing the global potential (Table 6.36; Table 6.44). <u>Food security</u>: Moderate benefits by helping farmers to harvest crops in advance of impending weather events or otherwise to make agricultural decisions to prepare for adverse events (Table 6.52; Fakhruddin et al. 2015; Genesio et al. 2011; Hillbruner and Moloney 2012).</p> |
| Risk sharing instruments | | | N D | | | <p><u>Mitigation</u>: Variable impacts- poor global coverage in the literature though studies from the US suggest a small increase in emissions from crop insurance and likely benefits from other risk sharing instruments (Table 6.20). <u>Adaptation</u>: Moderate benefits by buffering and transferring weather risk, saving farmers the cost of crop losses. However, overly subsidised insurance can undermine the market's role in pricing risks and thus depress more rapid adaptation strategies (Table 6.28; Meze-Hausken et al. 2009; Skees and Collier 2012; Jaworski 2016). <u>Desertification</u>: The impacts of risk sharing globally have not been quantified (Table 6.36). <u>Land degradation</u>: Variable impacts as evidence suggests that subsidised insurance in particular can increase crop production in marginal lands, and reforming this would lead to benefits (Table 6.44). <u>Food security</u>: Small to moderate benefits for food security, as risk sharing often promotes food supply sharing (Table 6.52).</p> |

1 Note: Cell colours correspond to the large, moderate and small categories shown in Table 6.53. Dark blue = large positive; mid-blue = moderate positive; light blue = small
2 positive; no colour = no effect; light red = small negative; mid-red = moderate negative; dark red = large negative; green = variable; ND = no data.

3

1 **6.4 Managing interactions and interlinkages**

2 Having assessed the potential of each response option for contributing to addressing mitigation,
3 adaptation, desertification, land degradation and food security in section 6.3, this section assesses the
4 feasibility of each response option with respect to cost, barriers, and issues of saturation and
5 reversibility (6.4.1), before assessing the sensitivity of the response options to future climate change
6 (6.4.2) and examining the contribution of each response option to ecosystem services (classified
7 according to Nature's Contribution to People (IPBES 2018) and to sustainable development (assessed
8 against the UN Sustainable Development Goals) (6.4.3). Section 6.4.4 examines opportunities for
9 implementation of integrated response options, paving the way to potential policies examined in
10 Chapter 7, before the consequences of delayed action are assessed in section 6.4.5.

11 **6.4.1 Feasibility of the integrated response options with respect to costs, barriers, 12 saturation and reversibility**

13 For each of the response options, Tables 6.62-6.69 summarise the feasibility with respect to saturation
14 and reversibility and cost, technological, institutional, socio-cultural and environmental and
15 geophysical barriers (the same barrier categories used in SR15).

16 Many land management options face issues of saturation and reversibility; however, these are not of
17 concern for the value chain and risk management options. Reversibility is an issue for all options that
18 increase terrestrial carbon stock, either through increased soil carbon or changes in land cover (e.g.,
19 reforestation, afforestation), since future changes in climate or land cover could result in reduced
20 carbon storage (Smith 2013). In addition, the benefits of options that improve land management (e.g.,
21 improved cropland management, improved grazing management) will cease if the practice is halted,
22 reversing any potential benefits.

23 The cost of the response options varies substantially, with some options having relatively low cost
24 (e.g., the cost of agroforestry is less than USD 10 tCO₂e⁻¹) while others have much higher costs (e.g.,
25 the cost of BECCS could be as much as USD 250 tCO₂e⁻¹). In addition to cost, other economic
26 barriers may prevent implementation; for example, agroforestry is a low- cost option (Smith et al.
27 2014a), but lack of reliable financial support could be a barrier (Hernandez-Morcillo et al. 2018).
28 Additionally, there are a number of reasons why even no cost options are not adopted, including risk
29 aversion, lack of information, market structure, externalities, and policies (Jaffe 2019).

30 Some of the response options have technological barriers that may limit their wide-scale application
31 in the near-term. For example, BECCS has only been implemented at small-scale demonstration
32 facilities (Kemper 2015a); challenges exist with upscaling these options to the levels discussed in this
33 Chapter.

34 Many response options have institutional and socio-cultural barriers. Institutional barriers include
35 governance, financial incentives and financial resources. For example, management of supply chains
36 faces challenges related to political will within trade regimes, economic laissez-faire policies that
37 discourage interventions in markets, and the difficulties of coordination across economic sectors
38 (Poulton et al. 2006; Cohen et al. 2009; Gilbert 2012a). Implementation of other options, e.g.,
39 BECCS, is limited by the absence of financial incentives.

40 Options like dietary change face socio-cultural barriers; while diets have changed in the past, they are
41 deeply culturally embedded and behaviour change is extremely difficult to effect, even when health
42 benefits are well known (Macdiarmid et al. 2018). For some options, the specific barrier is dependent
43 on the region. For example, barriers to reducing food waste in industrialised countries include
44 inconvenience, lack of financial incentives, lack of public awareness, and low prioritisation (Kummu
45 et al. 2012; Graham-Rowe et al. 2014). Barriers in developing countries include reliability of

- 1 transportation networks, market reliability, education, technology, capacity, and infrastructure
- 2 (Kummu et al. 2012).

1

2

3

Table 6.62 Feasibility of land management response options in agriculture, considering cost, technological, institutional, socio-cultural and environmental and geophysical barriers and saturation and reversibility. See also supplementary material.

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|--|------------|---------------|------|---------------|---------------|----------------|-------------------------------|---|
| Increased food productivity | | | | | | | | <u>Biophysical</u> : only if limited by climatic and environmental factors. <u>Sources</u> : Barnes and Thomson 2014; Martin et al. 2015; Olesen and Bindi 2002; Pretty and Bharucha 2014; Schut et al. 2016 |
| Improved cropland management | | | | | | | | <u>Institutional</u> : only in some regions (e.g., poor sustainability frameworks). <u>Sources</u> : Bryan et al. 2009b; Bustamante et al. 2014b; Madlener et al. 2006; Reichardt et al. 2009; Roesch-McNally et al. 2017; Singh and Verma 2007; Smith et al. 2008, 2014a |
| Improved grazing land management | | | | | | | | <u>Institutional</u> : only in some regions (e.g., need for extension services). <u>Sources</u> : Herrero et al. 2016; Singh and Verma 2007; Smith et al. 2008, 2015; McKinsey & Co., 2011; Ndoro et al., 2014; |
| Improved livestock management | | | | | | | | <u>Economic</u> : improved productivity is cost negative, but others (e.g. dietary additives) are expensive. <u>Institutional</u> : only in some regions (e.g. need for extension services). <u>Sources</u> : Herrero et al. 2016; McKinsey and Company 2009; Rojas-Downing et al. 2017b; Smith et al. 2008; Thornton et al. 2009; Beauchemin et al., 2008; Ndoro et al., 2014; |
| Agroforestry | | | | | | | | <u>Economic</u> : low cost but may lack reliable financial support. <u>Institutional</u> : only in some regions (e.g., seed availability). <u>Sources</u> : Lillesø et al. 2011; Meijer et al. 2015; Sileshi et al. 2008; Smith et al. 2007, 2014a |
| Agricultural diversification | | | | | | | | More support from extension services, access to inputs and markets, economic incentives for producing a certain crop or livestock product, research and investments focused on adapted varieties and climatic resilient systems, a combination of agricultural and non-agricultural activities (e.g., off farm jobs) are all important interventions aimed at overcoming barriers to agricultural diversification. <u>Sources</u> : Ahmed and Stepp 2016b; Barnes et al. 2015; Barnett and Palutikof 2015; Martin and Lorenzen 2016; Roesch-McNally et al. 2016; Waha et al. 2018 |
| Reduced grassland conversion to cropland | | | | | | | | <u>Economics</u> : Avoiding conversion is low cost, but there may be significant opportunity costs associated with foregone production of crops. <u>Institutional</u> : only in some regions (e.g., poor governance to prevent conversion) |
| Integrated water | | | | | | | | <u>Institutional</u> : effective implementation is dependent on the adoption of a combination of 'hard', |

| | | | | | | | | |
|------------|--|--|--|--|--|--|--|---|
| management | | | | | | | | infrastructural, and ‘soft’ institutional measures. <u>Socio-cultural</u> : Education can be a barrier and some strategies (e.g. site-specific water management, drip irrigation) can be expensive. Cultural / behavioural barriers are likely to be small. <u>Sources</u> : Dresner et al. 2015; Erwin 2009; Lotze et al. 2006; Thornton et al. 2009 |
|------------|--|--|--|--|--|--|--|---|

1 Note: For saturation and reversibility, a blue cell indicates that these issues are not important, and a red cell indicates that saturation and reversibility are concerns. For the
 2 cost column, a blue cell indicates low cost (< USD10 tCO₂e⁻¹ or < USD20 ha⁻¹), a yellow cell indicates medium cost (USD10-USD100 tCO₂e⁻¹ or USD20-USD100 ha⁻¹), and
 3 a red cell indicates high cost (>USD100 tCO₂e⁻¹ or USD200 ha⁻¹). The cost thresholds in USD tCO₂e⁻¹ are from Griscom et al. (2017a); thresholds in USD ha⁻¹ are chosen to
 4 be comparable, but precise conversions will depend on the response option. For the technological, institutional, socio-cultural and environmental and geophysical barriers,
 5 dark blue indicates high current feasibility (no barriers), mid-blue indicates medium current feasibility (moderate barriers) and light blue indicates low current feasibility
 6 (large barriers). Green represents variable barriers.

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|---------------------------------------|------------|---------------|------|---------------|---------------|----------------|-------------------------------|---|
| Improved forest management | | | | | | | | Sources: Seidl et al. 2014 |
| Reduced deforestation and degradation | | | | | | | | <u>Economic</u> : requires transaction and administration costs <u>Sources</u> : Kindermann et al. 2008; Overmars et al. 2014; Busch and Engelmann 2017; |
| Reforestation and forest restoration | | | | | | | | <u>Sources</u> : Strengers et al. 2008 |
| Afforestation | | | | | | | | <u>Sources</u> : Idris Medugu et al. 2010a; Kreidenweis et al. 2016b |

7 **Table 6.63 Feasibility of land management response options in forests, considering cost, technological, institutional, socio-cultural and environmental and**
 8 **geophysical barriers and saturation and reversibility. See also supplementary material.**

9 Note: See footnotes for Table 6.62.

10 **Table 6.64 Feasibility of land management response options for soils, considering cost, technological, institutional, socio-cultural and environmental and**
 11 **geophysical barriers and saturation and reversibility. See also supplementary material.**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|---------------------------------------|------------|---------------|------|---------------|---------------|----------------|-------------------------------|--|
| Increased soil organic carbon content | | | | | | | | <u>Institutional</u> : only in some regions (e.g., lack of institutional capacity). <u>Sources</u> : Smith et al. 2008; McKinsey and Company 2009; Baveye et al. 2018; Bustamante et al. 2014b; Reichardt et al. 2009; Smith 2006; Smith et al. 2007; Wollenberg et al. 2016 |
| Reduced soil erosion | | | | | | | | <u>Sources</u> : Haregeweyn et al. 2015 |
| Reduced soil salinisation | | | | | | | | Barriers depend on how salinisation and sodification are implemented. <u>Sources</u> : Bhattacharyya et al. 2015; CGIAR 2016; Dagar et al. 2016b; Evans and Sadler 2008; Greene et al. 2016; Machado and Serralheiro 2017 |
| Reduced soil compaction | | | | | | | | <u>Sources</u> : Antille et al. 2016; Chamen et al. 2015a |
| Biochar addition to soil | | | | | | | | Saturation and reversibility issues lower than for soil organic carbon. <u>Economics</u> : In general, biochar has high costs. However, a small amount of biochar potential could be available at negative cost, and some at low cost, depending on markets for the biochar as a soil amendment. <u>Institutional</u> : only in some regions (e.g., lack of quality standards). <u>Sources</u> : Chapter 4; Dickinson et al. 2014; Guo et al. 2016; Meyer et al. 2011; Shackley et al. 2011; Woolf et al. 2010 |

1 Note: See footnotes for Table 6.62.

2 **Table 6.65 Feasibility of land management response options in any/other ecosystems, considering cost, technological, institutional, socio-cultural and environmental**
3 **and geophysical barriers and saturation and reversibility. See also supplementary material.**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|------------------------|------------|---------------|------|---------------|---------------|----------------|-------------------------------|---|
| Fire management | | | | | | | | <u>Economic</u> : the cost of its implementation is moderate, since it requires constant maintenance, and can be excessive for some local communities. <u>Sources</u> : Freeman et al. 2017; Hurteau et al. 2014; North et al. 2015 |
| Reduced landslides and | | | | | | | | <u>Sources</u> : Gill and Malamud 2017; Maes et al. 2017; Noble et al. 2014 |

| | | | | | | | | |
|--|--|--|--|--|--|--|--|--|
| natural hazards | | | | | | | | |
| Reduced pollution including acidification | | | | | | | | Sources: Begum et al., 2011; Shah et al., 2018; Yamineva & Romppanen, 2017; WMO 2015 |
| Management of invasive species / encroachment | | | | | | | | <u>Technological</u> : in the case of natural enemies. <u>Socio-cultural</u> : Education can be a barrier, where populations are unaware of the damage caused by the invasive species, but cultural / behavioural barriers are likely to be small. <u>Institutional</u> : where agricultural extension and advice services are poorly developed. <u>Source</u> : Dresner et al. 2015 |
| Restoration and reduced conversion of coastal wetlands | | | | | | | | <u>Economic</u> : can be cost-effective at scale. <u>Institutional</u> : only in some regions (e.g., poor governance of wetland use). <u>Socio-cultural</u> : educational barriers (e.g., lack of knowledge of impact of wetland conversion), though cultural / behavioural barriers are likely to be small. <u>Sources</u> : Erwin 2009; Lotze et al. 2006 |
| Restoration and reduced conversion of peatlands | | | | | | | | <u>Institutional</u> : only in some regions (e.g., lack of inputs). <u>Sources</u> : Bonn et al. 2014; Worrall et al. 2009 |
| Biodiversity conservation | | | | | | | | <u>Economic</u> : While protected areas and other forms of biodiversity conservation can be cost-effective, they are often underfunded relative to needs. <u>Institutional</u> : There have been challenges in getting systematic conservation planning to happen, due to institutional fragmentation and overlapping mandates. <u>Socio-cultural</u> : Despite the fact that biodiversity conservation may provide co-benefits like water or carbon protection, local populations often have had social and cultural conflicts with protected areas and other forms of exclusionary biodiversity conservation that are imposed in a top-down fashion or which restrict livelihood options. <u>Sources</u> : Emerton et al. 2006; Hill et al. 2015; Langford et al. 2011; Larsen et al. 2012; Schleicher 2018; Wei et al. 2018; Wilkie et al. 2001 |

1 Note: See footnotes for Table 6.62.

2 **Table 6.66 Feasibility of land management response options specifically for CDR, considering cost, technological, institutional, socio-cultural and environmental**
 3 **and geophysical barriers and saturation and reversibility. See also supplementary material.**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|-----------------|------------|---------------|------|---------------|---------------|----------------|-------------------------------|--|
| Enhanced | | | | | | | | Permanence not an issue on the decadal timescales. <u>Institutional</u> : only in some regions (e.g., lack of infrastructure |

| | | | | | | | | |
|------------------------|--|--|--|--|--|--|--|--|
| weathering of minerals | | | | | | | | for this new technology). <u>Socio-cultural</u> : could occur in some regions, for example, due to minerals lying under undisturbed natural areas where mining might generate public acceptance issues. <u>Sources</u> : Renforth et al. 2012; Smith et al. 2016b; Taylor et al. 2016b |
| Bioenergy and BECCS | | | | | | | | <u>Economic</u> : while most estimates indicate the cost of BECCS as less than USD200 tCO ₂ ⁻¹ , there is significant uncertainty. <u>Technological</u> : while there are a few small BECCS demonstration facilities, BECCS has not been implemented at scale. <u>Sources</u> : IPCC SR15; Chapter 7; Kemper 2015; Sanchez and Kammen 2016; Vaughan and Gough 2016 |

1 Note: See footnotes for Table 6.62.

2 **Table 6.67 Feasibility of demand management response options, considering economic, technological, institutional, socio-cultural and environmental and**
 3 **geophysical barriers and saturation and reversibility. See also supplementary material.**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|---|-------------------|----------------------|-------------|----------------------|----------------------|-----------------------|--------------------------------------|--|
| Dietary change | | | | | | | | <u>Institutional</u> : only in some regions (e.g., poorly developed dietary health advice). <u>Sources</u> : Hearn et al. 1998; Lock et al. 2005; Macdiarmid et al. 2018; Wardle et al. 2000 |
| Reduced post-harvest losses | | | | | | | | |
| Reduced food waste (consumer or retailer) | | | | | | | | Specific barriers differ between developed and developing countries. <u>Sources</u> : Graham-Rowe et al. 2014; Kummu et al. 2012; Diaz-Ruiz et al. 2018; |
| Material substitution | | | | | | | | <u>Sources</u> : Gustavsson et al. 2006; Ramage et al. 2017 |

4 Note: See footnotes for Table 6.62.

5 **Table 6.68 Feasibility of supply management response options, considering cost, technological, institutional, socio-cultural and environmental and geophysical**
 6 **barriers and saturation and reversibility. See also supplementary material.**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|--|------------|---------------|------|---------------|---------------|----------------|-------------------------------|---|
| Sustainable sourcing | | | | | | | | <u>Economic</u> : the cost of certification and sustainable sourcing can lead to higher production costs. <u>Institutional</u> : there are some barriers to adopting sustainable sourcing in terms of getting governments on board with market-based policies. <u>Socio-cultural</u> : barriers include consumers unfamiliar with sustainably sourced goods. <u>Sources</u> : Capone et al. 2014; Ingram et al. 2016b |
| Management of supply chains | | | | | | | | <u>Economic</u> : Supply chain management and management of price volatility faces challenges from businesses in terms of economic costs of change. <u>Technological</u> : barriers like supply chain tracking. <u>Institutional</u> : barriers like political will against government action in markets. <u>Sources</u> : Cohen et al. 2009; Gilbert 2012; Poulton et al. 2006 |
| Enhanced urban food systems | | | | | | | | |
| Improved food processing and retailing | | | | | | | | <u>Economic</u> : The implementation of strategies to improve the efficiency and sustainability of retail and agri-food industries can be expensive. <u>Institutional</u> : Successful implementation is dependent on organisational capacity, the agility and flexibility of business strategies, the strengthening of public-private policies and effectiveness of supply-chain governance. |
| Improved energy use in food systems | | | | | | | | <u>Sources</u> : Baudron et al. 2015; Vlontzos et al. 2014 |

1 Note: See footnotes for Table 6.62.

2 **Table 6.69 Feasibility of risk management response options, considering cost, technological, institutional, socio-cultural and environmental and geophysical**
 3 **barriers and saturation and reversibility. See also supplementary material.**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical | Context and sources |
|----------------------------|------------|---------------|------|---------------|---------------|----------------|-------------------------------|--|
| Management of urban sprawl | | | | | | | | There are economic and political forces that benefit from less-regulated urban development. <u>Sources</u> : Tan et al. 2009 |

| | | | | | | | | |
|----------------------------|------|------|--------|-----------|------------|------------|-------|---|
| Livelihood diversification | Blue | Blue | White | White | White | Light Blue | White | <u>Economic</u> : Expanded diversification can cost additional financial resources. <u>Socio-cultural</u> : problems with adoption of new or unfamiliar crops and livelihoods. <u>Sources</u> : Ahmed and Stepp 2016b; Berman et al. 2012; Ngigi et al. 2017 |
| Use of local seeds | Blue | Blue | White | Dark Blue | Light Blue | Blue | White | <u>Economic</u> : Local seeds are highly cost effective, and do not require new technology. <u>Institutional</u> : barriers from agronomy departments and businesses promoting commercial seeds. <u>Socio-cultural</u> : preferences for some non-local seed sourced crops. <u>Sources</u> : Reisman 2017; Timmermann and Robaey 2016 |
| Disaster risk management | Blue | Blue | White | White | Light Blue | White | White | <u>Economic</u> : DRM systems can be initially costly, but usually pay for themselves over time. <u>Institutional</u> : some barriers in terms of getting initial support and will behind new systems. <u>Sources</u> : Birkmann et al. 2015b; Hallegatte 2012 |
| Risk sharing instruments | Blue | Blue | Yellow | White | White | Blue | White | There are few barriers to risk sharing instruments, as they are often low cost and low technology. <u>Socio-cultural</u> : some barriers to instruments like crop insurance, which some farmers in developing countries are not familiar with. <u>Sources</u> : Goodwin and Smith 2013 |

1 Note: See footnotes for Table 6.62.

2

1

2 **6.4.2 Sensitivity of the Integrated Response Options to climate change impacts**

3 With continued increases in warming, there are risks to the efficacy of some of the response options due
4 to future climate change impacts, such as increased climate variability and extreme events. While many of
5 the response options can help increase capacity to deliver adaptation benefits (section 6.3.2), beyond
6 certain thresholds of climate impacts they may be less effective or increasingly risky options. This
7 requires that some response options need to anticipate these climate impacts in their implementation. We
8 outline some of these impacts below.

9 *Agriculture response options:* Increased food productivity as a response option is highly sensitive to
10 climate change impacts. Chapter 5 (section 5.2.3.1) notes that global mean yields of some crops (maize
11 and soybean) decrease with warming, while others (rice and wheat) increase with warming, up to a
12 threshold of 3°C. Similarly, improved cropland management response options that rely on crop
13 diversification or improved varieties may face challenges in efficacy from production declines. Improved
14 grazing land management may continue to be feasible as a response option in the future under climate
15 change in northern regions but will likely become more difficult in tropical regions and Australia as
16 temperature rises will reduce the carrying capacity of lands (section 5.2.3.2; Nardone et al. 2010).
17 Improved livestock management also faces numerous challenges, particularly related to stresses on
18 animals from temperatures, water, and diseases; overall, livestock numbers are projected to decline 7.5–
19 9.6% by 2050 (section 5.2.3.2; Rivera-Ferre et al. 2016; Boone et al. 2018). Pastoralists may also be less
20 likely to implement improved measures due to other risks and vulnerabilities under climate change
21 (Thornton et al. 2009a).

22 The impact of climate change on agroforestry is more difficult to model than single crops in process-
23 based crop models, as agroforestry systems are far more complex (Luedeling et al. 2014); thus, it is
24 unknown how the efficacy of this response option might be impacted. Agricultural diversification has
25 been promoted as an adaptive strategy to climate impacts, given that diversity is known to increase
26 resiliency of agricultural and natural systems, such as in resistance to increased pests or diseases; it also
27 can provide diversified income portfolios when some crops may become sensitive to climate events
28 (Bradshaw et al. 2004; Lin 2011). Diversified farms are expected to increase in Africa by 2060 as
29 specialised farms with single crops face challenges under climate change (Seo 2010). However, it is not
30 known if these options and advantages of diversification have a temperature threshold beyond which they
31 are less effective.

32 Reduced grassland conversion is not likely to be affected as a response option *per se* since it is directed at
33 conserving natural grassland areas, but these areas may face increased pressures for conversion if farmers
34 experience crop failures under climate change and need to extensify holdings to make up for losses.
35 Lobell et al. (2013) have estimated the impacts of investment decisions to adapt to the effects of climate
36 change on crop yields to 2050 and find that cropland will expand over 23% more land area (over
37 3 Mkm²), mostly in Latin America and Sub-Saharan Africa.

38 Integrated water management to improve water availability and reliability of water for agricultural
39 production is likely to become more challenging in future scenarios of water declines, which are likely to
40 be regionally uneven (section 2.5, 6.4.4).

41 *Forest response options:* The availability of improved forest management as a response option can be
42 impacted by climate-induced changes, including increased diseases, pests and fires (Section 4.5.1.2; Dale
43 et al. 2001; Logan et al. 2003). These impacts will affect reforestation and afforestation response options

1 as well. Locatelli et al. (2015d) note that climate changes will influence seedling establishment, tree
2 growth and mortality, and the presence of invasive species and/or pests; these can be buffered with
3 modified silvicultural practices including species selection (Pawson et al. 2013). Climate changes can also
4 alter the sink capacity for vegetation carbon sequestration, reducing the potential for REDD, reforestation
5 and afforestation (Bonan 2008b; Mahli et al. 2002).

6 *Soil management:* Climate changes can alter the sink capacity for soil carbon sequestration, reducing the
7 potential for increased soil organic carbon as an option. Projected climate changes can reduce soil
8 resilience to extreme weather, pests and biological invasion, environmental pollutants and other pressures,
9 making reduced soil erosion and reduced soil compaction as response options harder to achieve (Smith et
10 al. 2015). Climate change will likely increase demand for irrigation in dryland areas, which can increase
11 risks of salinisation, diminishing the effectiveness of this response (Smith et al. 2015). Biochar additions
12 to soil may be affected by future climatic changes, such as rising soil temperatures, but little is known
13 given that most research on the subject is from laboratory and not *in situ* field experiments, and there are
14 wide estimates of the stability and residence times of biochar from this literature (Gurwick et al. 2013).

15 *Other ecosystem management:* Fire management is likely to become more challenging in a changing
16 climate; some studies suggest an 50% increase in fire occurrence by end of the century in circumboreal
17 forests (Flannigan et al. 2009). Landslide risks are related to climate through total rainfall, rainfall
18 intensity, air temperature and the general weather system (Gariano and Guzzetti 2016a); thus reduced
19 landslides and natural hazards as a response option will be made more difficult by increasing storms and
20 seasonality of rainfall events projected for many areas of the world. Reduced pollution is likely less
21 affected by climate change and can continue to be an option despite increasing temperatures.

22 Conversely, some invasive species may thrive under climate change, such as moving to new areas or
23 being less susceptible to control protocols (Hellmann et al. 2008). Conversion of coastal wetlands will be
24 more difficult to halt if loss of productive land elsewhere encourages development on these lands, but
25 coastal wetlands will likely adapt to increased CO₂ and higher sea levels through sediment accretion,
26 which will also enhance their capacity to act as carbon sinks (Duarte et al. 2013). While subarctic
27 peatlands are at risk due to warming, these are not the main peatlands that are at risk from agricultural
28 conversion (Tarnocai 2006); these peatlands, such as those in the tropics, may be more vulnerable in
29 hotter scenarios to water table alterations and fire risk (Gorham 1991). Biodiversity conservation, such as
30 through protected areas or corridors, may be threatened by increased land expansion under agriculture in
31 climate change scenarios, including the newly available land in northern climates that may become
32 agriculturally suited (Gimona et al. 2012), lessening the effectiveness of this response option.

33 *CDR:* The efficacy of enhanced weathering is not likely to be affected by future climate changes. On the
34 other hand, climate change will affect the productivity of bioenergy crops (Cronin et al. 2018),
35 influencing the mitigation potential of bioenergy and BECCS (Calvin et al. 2013a; Kyle et al. 2014).
36 There is uncertainty in the sign and magnitude of the effect of climate change on bioenergy crop yields.
37 As a result, there is uncertainty in whether climate change will increase or decrease the potential of
38 bioenergy and BECCS.

39 *Demand management of value chains:* For most response options in demand side management, the tools
40 are generally not made more difficult by future climate changes. For example, dietary change is not likely
41 to be affected by climate change, and in fact, the opposite is more likely; that diets will shift in response
42 to climate change impacts as reflected in high prices for some staple grains and meats, the productivity of
43 which may be reduced (Tigchelaar et al. 2018). However, there is some indication that fruit and vegetable
44 production will also be reduced in future scenarios, making healthier diets potentially harder to achieve in

1 some regions (Springmann et al. 2016). Reduced post-harvest losses and reduced food waste may become
2 an even more important option if water or heat stresses under climate change reduce overall harvests.
3 Material substitution does have risks related to the availability of products if there are declines in the
4 growth of forest and other biomass in certain future scenarios over time, although some evidence
5 indicates that biomass may increase in the short-term with limited warming (Boisvenue and Running
6 2006).

7 *Supply management of value chains:* Sustainable sourcing relies on being able to produce consumer
8 goods sustainably (palm oil, timber, cocoa, etc), and these may be at risk; for example, areas suitable for
9 oil palm production are estimated to decrease by 75% by 2100 (Paterson et al. 2017). Improved
10 management of supply chains is likely to increase in importance as a tool to manage food security, given
11 that climate change threatens to lead to more production shocks in the future (Baldos and Hertel 2015).
12 For enhanced urban food systems, climate stresses like heat island effects or increased water scarcity in
13 urban areas may reduce the viability of food production in certain urban systems (da Silva et al. 2012).
14 Improved food processing and retailing and improved energy use in agriculture are not likely to be
15 impacted by climate change.

16 *Risk management options:* Most risk management response options are not affected by climate impacts
17 *per se*, although the increased risks that people may face will increase the need for funding and support to
18 deploy these options. For example, disaster risk management will likely increase in importance in helping
19 people adapt to longer-term climate changes (Begum et al. 2014); it is also likely to cost more as
20 increased impacts of climate change, such as intensification or frequency of storm events may increase.
21 Management of urban sprawl may also be challenged by increased migration driven by climate change, as
22 people displaced by climate change may move to unregulated urban areas (Adamo 2010). Livelihood
23 diversification can assist in adapting to climate changes and is not likely to be constrained as a response
24 option, as climate-sensitive livelihoods may be replaced by others less so. Use of local seeds as an
25 effective response options may depend on the specific types of seeds and crops used, as some may not be
26 good choices under increased heat and water stress (Gross et al. 2017). Risk sharing instruments are
27 unlikely to be affected by climate change, with the exception of index and crop insurance, which may
28 become unaffordable if too many climate shocks result in insurance claims decreasing the ability of the
29 industry to provide this tool (Mills 2005).

30

31

32 **Cross-Chapter Box 8: Ecosystem services and Nature's Contributions to** 33 **People, and their relation to the land-climate system**

34 Pamela McElwee (The United States of America), Jagdish Krishnaswamy (India), Lindsay Stringer
35 (United Kingdom)

36 This Cross-Chapter Box describes the concepts of *ecosystem services (ES)* and *nature's contributions to*
37 *people (NCP)*, and their importance to climate-land interactions. ES have become a useful concept to
38 describe the benefits that humans obtain from ecosystems and have strong relevance to sustainable land
39 management (SLM) decisions and their outcomes, while NCP is a new approach championed by the
40 Intergovernmental Panel on Biodiversity and Ecosystem Services (IPBES) (explained below). It is timely
41 that this SRCCL report includes attention to ES/NCP, as the previous Special Report on Land-Use, Land-
42 Use Change and Forestry (LULUCF) did not make use of these concepts and focused mostly on carbon

1 fluxes in land-climate interactions (IPCC 2000). The broader mandate of SRCCL is to address not just
2 climate but land degradation, desertification and food security issues, all of which are closely linked to the
3 provisioning of various ES/NCP, and the Decision and Outline for SRCCL explicitly requests an
4 examination of how desertification and degradation “impacts on ecosystem services (e.g. water, soil and
5 soil carbon and biodiversity that underpins them)”. Attention to ES/NCP is particularly important in
6 discussing co-benefits, trade-offs and adverse side effects of potential climate change mitigation, land
7 management, or food security response options, as many actions may have positive impacts on climate
8 mitigation or food production but may also come with a decline in ES provisioning, or adversely impact
9 biodiversity {see 6.4.3}. This box considers the importance of the ES/NCP concepts, how definitions
10 have changed over time, continuing debates over operationalisation and use of these ideas, and finally
11 concludes with how ES/NCP are treated in various chapters in this report.

12 While the first uses of the term “ecosystem services” appeared in the 1980s (Lele et al. 2013; Mooney and
13 Ehrlich 1997), the roots of interest in ES extends back to the late 1960s and the extinction crisis, with
14 concern that species decline might cause loss of valuable benefits to humankind (King 1966; Helliwell
15 1969; Westman 1977). While concern over extinction was explicitly linked to biodiversity loss, later ideas
16 beyond biodiversity have animated interest in ES, including the multi-functional nature of ecosystems. A
17 seminal paper by Costanza et al. (1997) attempted to put an economic value on the stocks of global ES
18 and natural capital on which humanity relied. Attention to ES expanded rapidly after the Millennium
19 Ecosystem Assessment (Millenium Ecosystem Assessment (MA) 2005), and the linkages between ES and
20 economic valuation of these functions were addressed by the Economics of Ecosystems and Biodiversity
21 study (TEEB 2009). The ES approach has increasingly been used in global and national environmental
22 assessments, including the United Kingdom National Ecosystem Assessment (Watson et al. 2011), and
23 recent and ongoing regional and global assessments organised by the Intergovernmental Science-Policy
24 Platform on Biodiversity and Ecosystem Services (IPBES) (Díaz et al. 2015). IPBES has recently
25 completed an assessment on land degradation and restoration that addresses a range of ES issues of
26 relevance to the SRCCL report (IPBES 2018).

27 The MA defined ES as “the benefits that ecosystems provide to people,” and identified four broad
28 groupings of ES: *provisioning services* such as food, water, or timber; *regulating services* that have
29 impacts on climate, diseases or water quality, among others; *cultural services* that provide recreational,
30 aesthetic, and spiritual benefits; and *supporting services* such as soil formation, photosynthesis, and
31 nutrient cycling (Millenium Ecosystem Assessment (MA) 2005). The MA emphasised that people are
32 components of ecosystems engaged in dynamic interactions, and particularly assessed how changes in ES
33 might impact human well-being, such as access to basic materials for living (shelter, clothing, energy);
34 health (clean air and water); social relations (including community cohesion); security (freedom from
35 natural disasters); and freedom of choice (the opportunity to achieve) (Millenium Ecosystem Assessment
36 (MA) 2005). Upon publication of the MA, incorporation of ES into land use change assessments
37 increased dramatically, including studies on how to maximise provisioning of ES alongside human well-
38 being (Carpenter et al. 2009); how intensive food production to feed growing populations required trading
39 off a number of important ES (Foley et al. 2005); and how including ES in GCMs indicated increasing
40 vulnerability to ES change or loss in future climate scenarios (Schröter et al. 2005).

41 Starting in 2015, IPBES has introduced a new related concept to ES, that of *nature’s contributions to*
42 *people (NCP)*, which are defined as “all the contributions, both positive and negative, of living nature
43 (i.e., diversity of organisms, ecosystems and their associated ecological and evolutionary processes) to the
44 quality of life of people” (Díaz et al. 2018). NCP are divided into regulating NCP, non-material NCP, and
45 material NCP, a different approach than used by the MA (see figure 1). However, IPBES has stressed

NCP are a particular way to think of ES, rather than a replacement for ES. Rather, the concept of NCP is proposed to be broader umbrella to engage a wider range of scholarship, particularly from the social sciences and humanities, and a wider range of values, from intrinsic to instrumental to relational, particularly those held by indigenous and other peoples (Redford and Adams 2009; Schröter et al. 2014; Pascual et al. 2017; Díaz et al. 2018). The differences between the MA and IPBES approaches can be seen in Table 1.

Table 1. Comparison of MA and IPBES categories and types of ES and NCP

| MA category | MA: Ecosystem Services | IPBES category | IPBES: Nature's Contributions to People |
|--|---|---------------------------------|--|
| Supporting services | Soil formation | | |
| | Nutrient cycling | | |
| | Primary production | | |
| Regulating services | | Regulating Contributions | Habitat creation and maintenance |
| | Pollination | | Pollination and dispersal of seeds and other propagules |
| | Air quality regulation | | Regulation of air quality |
| | Climate regulation | | Regulation of climate |
| | Water regulation | | Regulation of ocean acidification |
| | See above | | Regulation of freshwater quantity, flow and timing |
| | Water purification and waste treatment | | Regulation of freshwater and coastal water quality |
| | Erosion regulation | | Formation, protection and decontamination of soils and sediments |
| Natural hazard regulation | Regulation of hazards and extreme events | | |
| Pest regulation and disease regulation | Regulation of organisms detrimental to humans | | |
| Provisioning Services | Fresh water | Material Contributions | Energy |
| | Food | | Food and feed |
| | Fibre | | Materials and assistance |

| | | | |
|--------------------------|---------------------------------------|----------------------------------|--|
| | Medicinal and biochemical and genetic | | Medicinal, biochemical and genetic resources |
| Cultural Services | Aesthetic values | Nonmaterial Contributions | Learning and inspiration |
| | Recreation and ecotourism | | Physical and psychological experiences |
| | Spiritual and religious values | | Supporting identities |
| | | | Maintenance of options |

1 Sources: Millenium Ecosystem Assessment (MA) 2005; Díaz et al. 2018

2 While there are many similarities between ES and NCP as seen above, the IPBES decision to use the NCP
3 concept has been controversial, with some people arguing that an additional term is superfluous, that it
4 incorrectly associates ES with economic valuation, and that the NCP concept is not useful for policy
5 uptake (Braat 2018; Peterson et al. 2018). Others have argued that the MA approach is outdated, did not
6 explicitly address biodiversity, and confused different concepts, like economic goods, ecosystem
7 functions, and general benefits (Boyd and Banzhaf 2007). Moreover, for both ES and NCP approaches, it
8 has been difficult to make complex ecological processes and functions amenable to assessments that can
9 be used and compared across wider landscapes, different policy actors, and multiple stakeholders (de
10 Groot et al. 2002; Naem et al. 2015; Seppelt et al. 2011). There remain competing categorisation
11 schemes for ES, as well as competing metrics on how most ES might be measured (Wallace 2007;
12 Potschin and Haines-Young 2011; Danley and Widmark 2016; Nahlik et al. 2012). The implications of
13 these discussions for this SRCCL report is that there remain many areas of uncertainty with regard to
14 much ES/NCP measurement and valuation, which will have ramifications for choosing response options
15 and policies.

16 This report addresses ES/NCP in multiple ways. Individual chapters have used the term ES in most cases,
17 especially since the preponderance of existing literature uses the ES terminology. For example, Chapter 2
18 discusses CO₂ fluxes, nutrients, and water budgets as important ES deriving from land-climate
19 interactions. Chapters 3 and 4 discuss issues such as biomass production, soil erosion, biodiversity loss,
20 and other ES affected by land use change. Chapter 5 discusses both ES and NCP issues surrounding food
21 system provisioning and trade-offs.

22 In chapter 6, the concept of NCP is used. For example, in chapter 6 Tables 6.70 to 6.72, possible response
23 options to respond to climate change, to address land degradation or desertification, and to ensure food
24 security are cross-referenced against the 18 NCP identified by Díaz et al. (2018) to see where there are
25 co-benefits and adverse side-effects. For instance, while BECCS may deliver on climate mitigation, it
26 results in a number of adverse side-effects that are significant with regard to water provisioning, food and
27 feed availability, and loss of supporting identities if BECCS competes against local land uses of cultural
28 importance. Chapter 7 has an explicit section 7.2.2.2 that covers risks due to loss of biodiversity and ES
29 and Table 7.1 that includes policy responses to various land-climate-society hazards, some of which are
30 likely to enhance risk of loss of biodiversity and ES. A case-study on the impact of renewable energy on
31 biodiversity and ES is also included. Chapter 7 also notes that because there is no SDG covering fresh-

1 water biodiversity and aquatic ecosystems; this policy gap may have adverse consequences for the future
2 of rivers and associated ES.

4 **6.4.3 Impacts of integrated response options on Nature’s Contributions to People and the** 5 **UN Sustainable Development Goals**

6 In addition to evaluating the importance of our response options for climate mitigation, adaptation, land
7 degradation, desertification and food security, it is also necessary to pay attention to other co-benefits and
8 trade-offs that may be associated with these responses. How the different options impact progress toward
9 the SDG can be a useful shorthand for looking at the social impacts of these response options. Similarly,
10 looking at how these response options increase or decrease the supply of ecosystem services/NCP (see
11 Cross-Chapter Box 8 on Ecosystem Services in this chapter) can be a useful shorthand for a more
12 comprehensive environmental impact beyond climate and land. Such evaluations are important as
13 response option may lead to unexpected trade-offs with social goals (or potential co-benefits) and impacts
14 on important environmental indicators like water or biodiversity. Similarly, there may be important
15 synergies and co-benefits associated with some response options that may increase their cost-
16 effectiveness or attractiveness. As we note in section 6.4.4, many of these synergies are not automatic,
17 and are dependent on well-implemented and coordinated activities in appropriate environmental contexts
18 (6.4.4.1), often requiring institutional and enabling conditions for success and participation of multiple
19 stakeholders (6.4.4.3).

20 In the following sections and tables, we evaluate each response option against 17 SDG and 18 NCP.
21 Some of the SDG categories appear similar to each other, such as SDG 13 on “climate action” and an
22 NCP titled “climate regulation”. However, SDG 13 includes targets for both mitigation and adaptation, so
23 options were weighed by whether they were useful for one or both. On the other hand, the NCP
24 “regulation of climate” does not include an adaptation component, and refers to specifically to “positive
25 or negative effects on emissions of greenhouse gases and positive or negative effects on biophysical
26 feedbacks from vegetation cover to atmosphere, such as those involving albedo, surface roughness, long-
27 wave radiation, evapotranspiration (including moisture-recycling) and cloud formation or direct and
28 indirect processes involving biogenic volatile organic compounds (BVOC), and regulation of aerosols and
29 aerosol precursors by terrestrial plants and phytoplankton” (Díaz et al. 2018).

30 In all tables, colours represent the direction of impact: positive (blue) or negative (brown), and the scale
31 of the impact (dark colours for large impact and/or strong evidence to light colours for small impact
32 and/or less certain evidence). Supplementary tables show the values and references used to define the
33 colour coding used in all tables. In cases where there is no evidence of an interaction or at least no
34 literature on such interactions, the cell is left blank. In cases where there are both positive and negative
35 interactions and the literature is uncertain about the overall impact, a note appears in the box. In all cases,
36 many of these interactions are contextual, or the literature only refers to certain co-benefits in specific
37 regions or ecosystems, so readers are urged to consult the supplementary tables for the specific caveats
38 that may apply.

39 **6.4.3.1 Impacts of integrated response options on Nature’s Contributions to People**

40 Tables 6.70–6.72 summarise the impacts of the response options on NCP supply. Examples of synergies
41 between response options and NCP include positive impacts on habitat maintenance (NCP 1) from
42 activities like invasive species management and agricultural diversification. For the evaluation process,
43 we considered that NCP are about ecosystems, therefore options which may have overall positive effects,

1 but which are *not* ecosystem-based are not included; for example, improved food transport and
2 distribution could reduce ground-level ozone and thus improve air quality, but this is not an ecosystem-
3 based NCP. Similarly, energy efficiency measures would increase energy availability, but the ‘energy’
4 NCP refers specifically to biomass-based fuel provisioning. This necessarily means that the land
5 management options have more direct NCP effects than the value chain or governance options, which are
6 less ecosystem-focused.

7 In evaluating NCP, we have also tried to avoid ‘indirect’ effects – that is a response option might increase
8 household income which then could be invested in habitat-saving actions, or dietary change would lead to
9 conservation of natural areas, which would then led to increased water quality. Similarly, material
10 substitution would increase wood demand, which in turn might lead to deforestation which might have
11 water regulation effects. These can all be considered *indirect* impacts on NCP, which were not evaluated⁷.
12 Instead, the assessment focuses as much as possible on *direct* effects only: for example, local seeds
13 policies preserve local landraces, which *directly* contribute to ‘maintenance of genetic options’ for the
14 future. Therefore, this NCP table is a conservative estimation of NCP effects; there are likely many more
15 secondary effects, but they are too difficult to assess, or the literature is not yet complete or conclusive.

16 Further, many NCP trade-off with one another (Rodriguez et al 2006), so supply of one might lead to less
17 availability of another – for example, use of ecosystems to produce bioenergy will likely lead to decreases
18 in water availability if mono-cropped high intensity plantations are used (Gasparaos et al 2011).

19 Overall, several response options stand out as having co-benefits across 10 or more NCP with no adverse
20 impacts: improved cropland management, agroforestry, forest management and forest restoration,
21 increased soil organic content, fire management, restoration and avoided conversion of coastal wetlands,
22 and use of local seeds. Other response options may have strengths in some NCP but require trade-offs
23 with others. For example, reforestation and afforestation bring many positive benefits for climate and
24 water quality but may trade-off with food production (Table 6.70). Several response options, including
25 increased food productivity, bioenergy and BECCS, and some risk sharing instruments like crop
26 insurance, have significant negative consequences across multiple NCP.

⁷ FOOTNOTE: The exception is NCP 6, regulation of ocean acidification, which is by itself an indirect impact. Any option that sequesters CO₂ would lower the atmospheric CO₂ concentration, which then indirectly increases the seawater pH. Therefore, any action that directly increases the amount of sequestered carbon is noted in this column, but not any action that avoids land use change and therefore indirectly avoids CO₂ emissions.

Table 6.70 Impacts on Nature’s Contributions to People of integrated response options based on land management

| <u>Integrated response options based on land management</u> | Habitat creation and maintenance | Pollination and dispersal of seeds and other propagules | Regulation of air quality | Regulation of climate | Regulation of ocean acidification | Regulation of freshwater quantity, flow and timing | Regulation of freshwater and coastal water quality | Formation, protection and decontamination of soils and sediments | Regulation of hazards and extreme events | Regulation of organisms detrimental to humans | Energy | Food and feed | Materials and assistance | Medicinal, biochemical and genetic resources | Learning and inspiration | Physical and psychological experiences | Supporting identities | Maintenance of options |
|---|----------------------------------|---|---------------------------|-----------------------|-----------------------------------|--|--|--|--|---|--------|---------------|--------------------------|--|--------------------------|--|-----------------------|------------------------|
| Increased food productivity | | | | | | | | | | | | | | | | | | |
| Improved cropland management | | | | | | | | | | | | | | | | | | |
| Improved grazing land management | | | | | | | | | | | | | | | | | | |
| Improved livestock management | | | | | | | | | | | | | | | | | | |
| Agroforestry | | | | | | | | | | | | | | | | | | |
| Agricultural diversification | | | | | | | | | | | | | | | | | | |
| Avoidance of conversion of grassland to cropland | | | | | | | | | | | | | | | | | | |
| Integrated water management | | | | | | | | | | | | | + | or - | | | | |
| Forest management and forest restoration | | | | | | | | | + | or - | | | + | or - | | | | |

| | | | | | | | | | | | | | | | | | |
|---------------------------------------|--|--|--|--|--|--|--------|--------|--|--|--|--|--|--|--|--|--|
| Reduced deforestation and degradation | | | | | | | | | | | | | | | | | |
| Reforestation | | | | | | | | + or - | | | | | | | | | |
| Afforestation | | | | | | | + or - | + or - | | | | | | | | | |

| | | | | | | | | | | | | | | | | | |
|---------------------------------------|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|
| Increased soil organic carbon content | | | | | | | | | | | | | | | | | |
| Reduced soil erosion | | | | | | | | | | | | | | | | | |
| Reduced soil salinisation | | | | | | | | | | | | | | | | | |
| Reduced soil compaction | | | | | | | | | | | | | | | | | |
| Biochar addition to soil | | | | | | | | | | | | | | | | | |

| | | | | | | | | | | | | | | | | | |
|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|
| Fire management | | | | | | | | | | | | | | | | | |
| Reduced landslides and natural hazards | | | | | | | | | | | | | | | | | |
| Reduced pollution including acidification | | | | | | | | | | | | | | | | | |
| Management of invasive species / encroachment | | | | | | | | | | | | | | | | | |
| Restoration and avoided conversion of coastal wetlands | | | | | | | | | | | | | | | | | |
| Restoration and avoided conversion of peatlands | | | | | | | | | | | | | | | | | |
| Biodiversity conservation | | | | | | | | | | | | | | | | | |

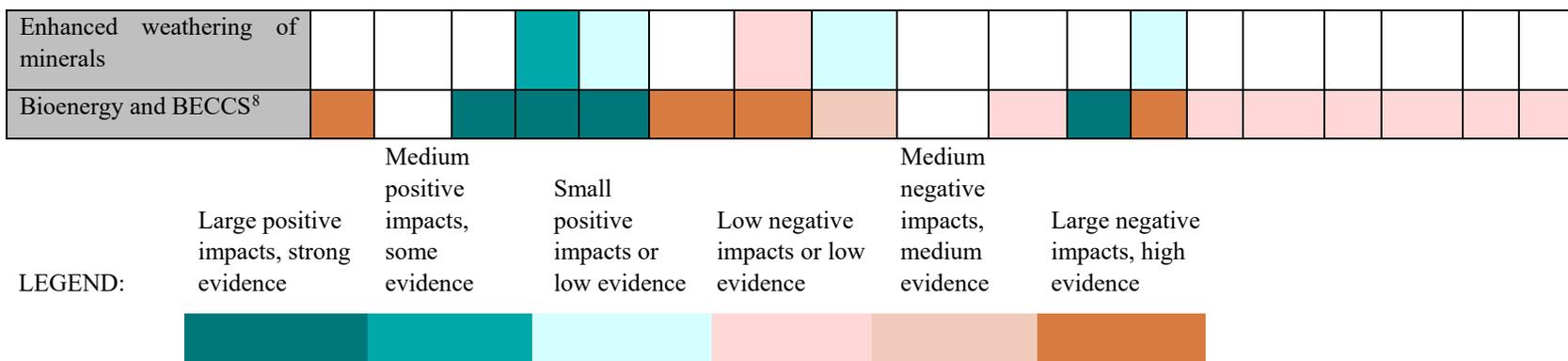


Table 6.71 Impacts on Nature’s Contributions to People of integrated response options based on value chain management

| | Habitat creation and maintenance | Pollination and dispersal of seeds and other propagules | Regulation of air quality | Regulation of climate | Regulation of ocean acidification | Regulation of freshwater quantity, flow and timing | Regulation of freshwater and coastal water quality | Formation, protection and decontamination of soils and | Regulation of hazards and extreme events | Regulation of organisms detrimental to humans | Energy | Food and feed | Materials and assistance | Medicinal, biochemical and genetic resources | Learning and inspiration | Physical and psychological experiences | Supporting identities | Maintenance of options |
|--|----------------------------------|---|---------------------------|-----------------------|-----------------------------------|--|--|--|--|---|--------|---------------|--------------------------|--|--------------------------|--|-----------------------|------------------------|
| <u>Integrated response options based on value chain management</u> | | | | | | | | | | | | | | | | | | |
| Dietary change | | | | | | | | | | | | | | | | | | |
| Reduced post-harvest losses | | | | | | | | | | | | | | | | | | |

⁸ FOOTNOTE: Note that this refers to large areas of bioenergy crops capable of producing large mitigation benefits (> 3 GtCO₂ yr⁻¹). The effect of bioenergy and BECCS on NCPs is scale and context dependent (see Cross-Chapter Box 7 in this chapter; Section 6.2).

| | | | | | | | | | | | | | | | | | |
|---|---|--|--|--|--|--|--------------------------------------|--|--|--|--|--|--|--|--|--|--|
| Reduced food waste (consumer or retailer) | Large positive impacts, strong evidence | | | Medium positive impacts, some evidence | | Small positive impacts or low evidence | Low negative impacts or low evidence | | | | | Medium negative impacts, medium evidence | | | | | |
| Material substitution | Large negative impacts, high evidence | | | Small positive impacts or low evidence | | | | | | | | | Medium negative impacts, medium evidence | | | | |

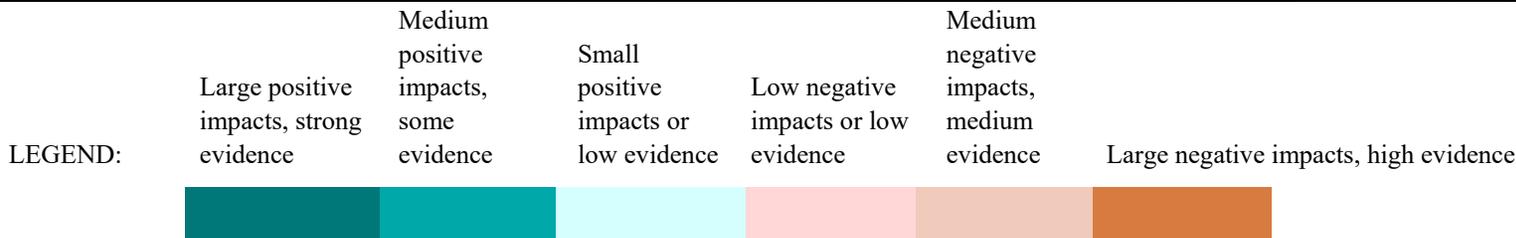
| | | | | | | | | | | | | | | | | | | |
|-------------------------------------|--|--|--|--|--|--|--------------------------------------|--|--|--|--|--|--|--|--|--|--|--|
| Sustainable sourcing | Small positive impacts or low evidence | | | Medium positive impacts, some evidence | | Small positive impacts or low evidence | Low negative impacts or low evidence | | | | | Small positive impacts or low evidence | Medium positive impacts, some evidence | Medium negative impacts, medium evidence | Small positive impacts or low evidence | | | |
| Management of supply chains | | | | | | | | | | | | | Medium positive impacts, some evidence | Small positive impacts or low evidence | | | | |
| Enhanced urban food systems | Small positive impacts or low evidence | | | | | | | | | Medium negative impacts, medium evidence | | Low negative impacts or low evidence | | Small positive impacts or low evidence | Small positive impacts or low evidence |
| Improved food processing and retail | | | | | | | | | | | | | | | | | | |
| Improved energy use in food systems | | | | | | | | | | | | | | | | | | |

LEGEND:

| | | | | | |
|--|--|--|---|--|--|
| Large positive impacts, strong evidence | Medium positive impacts, some evidence | Small positive impacts or low evidence | Low negative impacts or low evidence | Medium negative impacts, medium evidence | Large negative impacts, high evidence |
|  |  |  |  |  |  |

Table 6.72 Impacts on Nature’s Contributions to People of integrated response options based on risk management

| <u>Integrated response options based on risk management</u> | Habitat creation and maintenance | Pollination and dispersal of seeds and other propagules | Regulation of air quality | Regulation of climate | Regulation of ocean acidification | Regulation of freshwater quantity, flow and timing | Regulation of freshwater and coastal water quality | Formation, protection and decontamination of soils and sediments | Regulation of hazards and extreme events | Regulation of organisms detrimental to humans | Energy | Food and feed | Materials and assistance | Medicinal, biochemical and genetic resources | Learning and inspiration | Physical and psychological experiences | Supporting identities | Maintenance of options | |
|---|---|---|--|--------------------------------------|--|--|--|--|--|---|---------------------------------------|---------------------------------------|---------------------------------------|--|---------------------------------------|--|---------------------------------------|---------------------------------------|---------------------------------------|
| Management of urban sprawl | Large positive impacts, strong evidence | Medium positive impacts, some evidence | Small positive impacts or low evidence | Low negative impacts or low evidence | Medium negative impacts, medium evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence |
| Livelihood diversification | Large positive impacts, strong evidence | Medium positive impacts, some evidence | Small positive impacts or low evidence | Low negative impacts or low evidence | Medium negative impacts, medium evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence |
| Use of local seeds | Large positive impacts, strong evidence | Medium positive impacts, some evidence | Small positive impacts or low evidence | Low negative impacts or low evidence | Medium negative impacts, medium evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence |
| Disaster risk management | Large positive impacts, strong evidence | Medium positive impacts, some evidence | Small positive impacts or low evidence | Low negative impacts or low evidence | Medium negative impacts, medium evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence |
| Risk sharing instruments | Large positive impacts, strong evidence | Medium positive impacts, some evidence | Small positive impacts or low evidence | Low negative impacts or low evidence | Medium negative impacts, medium evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence | Large negative impacts, high evidence |



6.4.3.2 Impacts of integrated response options on the UN Sustainable Development Goals

Tables 6.73–6.75 summarise the impact of the integrated response options on the UN SDG. Some of the synergies between response options and SDG in the literature include positive poverty reduction impacts (SDG 1) from activities like improved water management or improved management of supply chains, or positive gender impacts (SDG 5) from livelihood diversification or use of local seeds. Because many land management options only produce indirect or unclear effects on SDG, we did not include these where there was no literature. Therefore, the value chain and governance options appear to offer more direct benefits for SDG.

However, it is noted that some SDG are internally difficult to assess because they contain many targets, not all of which could be evaluated (e.g., SDG 17 is about partnerships, but has targets ranging from foreign aid to debt restructuring to technology transfer to trade openness). Additionally, it is noted that some SDG contradict one another – for example, SDG 9 to increase industrialisation and infrastructure and SDG 15 to improve life on land. More industrialisation is likely to lead to increased resource demands with negative effects on habitats. Therefore, a positive association on one SDG measure might be directly correlated with a negative measure on another, and the table needs to be read with caution for that reason. The specific caveats on each of these interactions can be found in the supplementary material tables in the Chapter 6 appendix.

Overall, several response options have co-benefits across 10 or more SDG with no adverse side effects on any SDG: increased food production, improved grazing land management, agroforestry, integrated water management, reduced post-harvest losses, sustainable sourcing, livelihood diversification and disaster risk management. Other response options may have strengths in some SDG but require trade-offs with others. For example, use of local seeds bring many positive benefits for poverty and hunger reduction, but may reduce international trade (SDG 17). Other response options like enhanced urban food systems, management of urban sprawl, or management of supply chains are generally positive for many SDG but may trade-off with one, like clean water (SDG 6) or decent work (SDG 8), as they may increase water use or slow economic growth. Several response options, including avoidance of grassland conversion, reduced deforestation and degradation, reforestation and afforestation, biochar, restoration and avoided conversion of peatlands and coastlands, have trade-offs across multiple SDG, primarily as they prioritise land health over food production and poverty reduction. Several response options such as bioenergy and BECCS and some risk sharing instruments, such as crop insurance, trade-off over multiple SDG with potentially significant adverse consequences.

Overall, across both categories of both SDG and NCP, 17 of 40 options deliver co-benefits or no adverse side-effects for the full range of NCP and SDG. This include most agriculture- and soil-based land management options, many ecosystem-based land management options, improved forest management, reduced post-harvest losses, sustainable sourcing, improved energy use in food systems, and livelihood diversification. Only three options (afforestation, bioenergy and BECCS and some types of risk sharing instruments, such as crop insurance) have potentially adverse side-effects for five or more NCP or SDG.

Table 6.73 Impacts on the UN SDG of integrated response options based on land management

| <u>Integrated response options based on land management</u> | GOAL 1: No Poverty | GOAL 2: Zero Hunger | GOAL 3: Good Health and Well-being | GOAL 4: Quality Education | GOAL 5: Gender Equality | GOAL 6: Clean Water and Sanitation | GOAL 7: Affordable and Clean Energy | GOAL 8: Decent Work and Economic Growth | GOAL 9: Industry, Innovation and Infrastructure | GOAL 10: Reduced Inequality | GOAL 11: Sustainable Cities and Communities | GOAL 12: Responsible Consumption and Production | GOAL 13: Climate Action | GOAL 14: Life Below Water | GOAL 15: Life on Land | GOAL 16: Peace and Justice Strong Institutions | GOAL 17: Partnerships to achieve the Goal |
|---|--------------------|---------------------|------------------------------------|---------------------------|-------------------------|------------------------------------|-------------------------------------|---|---|-----------------------------|---|---|-------------------------|---------------------------|-----------------------|--|---|
| Increased food productivity | | | | | | | | | | | | | | | | | |
| Improved cropland management | | | | | | | | | | | | | | | | | |
| Improved grazing land management | | | | | | | | | | | | | | | | | |
| Improved livestock management | | | | | | | | | | | | | | | | | |
| Agroforestry | | | | | | | | | | | | | | | | | |
| Agricultural diversification | | | | | | | | | | + | | | | | | | |
| Avoidance of conversion of grassland to cropland | | | | | | | | | | | | | | | | | |
| Integrated water management | | | | | | | | | | | | | | | | | |
| Forest management and forest restoration | | | | | | | | | | | | | | | | | |
| Reduced deforestation and degradation | + or - | | | | | | | | | | | | | | | | |

| | | | | | | | | | | | | | | | | | | |
|--|--------|--------|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|--|
| Reforestation | + or - | | | | | | | | | | | | | | | | | |
| Afforestation | | | | | | | | | | | | | | | | | | |
| Increased soil organic carbon content | | | | | | | | | | | | | | | | | | |
| Reduced soil erosion | | | | | | | | | | | | | | | | | | |
| Reduced soil salinisation | | | | | | | | | | | | | | | | | | |
| Reduced soil compaction | | | | | | | | | | | | | | | | | | |
| Biochar addition to soil | | | | | | | | | | | | | | | | | | |
| Fire management | | | | | | | | | | | | | | | | | | |
| Reduced landslides and natural hazards | | | | | | | | | | | | | | | | | | |
| Reduced pollution including acidification | | | | | | | | | | | | | | | | | | |
| Management of invasive species / encroachment | | | | | | | | | | | | | | | | | | |
| Restoration and avoided conversion of coastal wetlands | + or - | + or - | | | | | | | | | | | | | | | | |
| Restoration and avoided conversion of peatlands | | | | | | | | | | | | | | | | | | |
| Biodiversity conservation | | | | | | | | | | | | | | | | | | |
| Enhanced weathering of minerals | | | | | | | | | | | | | | | | | | |

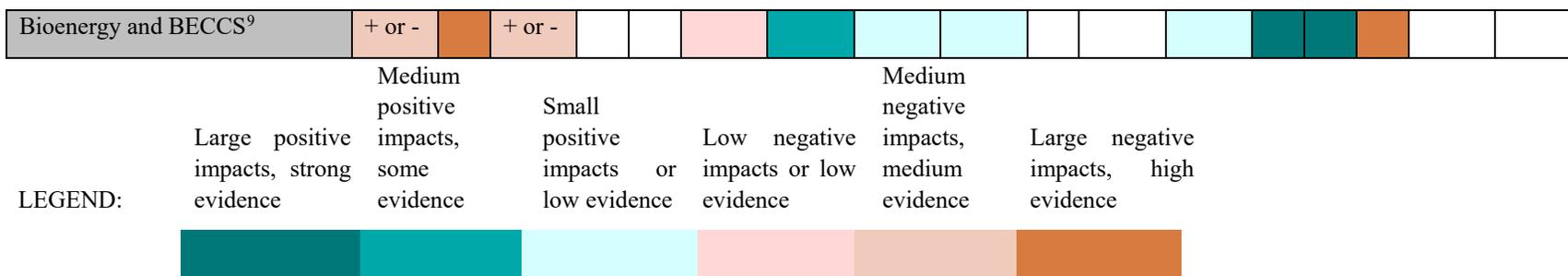


Table 6.74 Impacts on the UN SDG of integrated response options based on value chain interventions

| <u>Integrated response options based on value chain management</u> | GOAL 1: No Poverty | GOAL 2: Zero Hunger | GOAL 3: Good Health and Well-being | GOAL 4: Quality Education | GOAL 5: Gender Equality | GOAL 6: Clean Water and Sanitation | GOAL 7: Affordable and Clean Energy | GOAL 8: Decent Work and Economic Growth | GOAL 9: Industry, Innovation and Infrastructure | GOAL 10: Reduced Inequality | GOAL 11: Sustainable Cities and Communities | GOAL 12: Responsible Consumption and Production | GOAL 13: Climate Action | GOAL 14: Life Below Water | GOAL 15: Life on Land | GOAL 16: Peace and Justice Strong Institutions | GOAL 17: Partnerships to achieve the Goal |
|--|--------------------|---------------------|------------------------------------|---------------------------|-------------------------|------------------------------------|-------------------------------------|---|---|-----------------------------|---|---|-------------------------|---------------------------|-----------------------|--|---|
| Dietary change | Light Blue | Dark Blue | Dark Blue | | | Dark Blue | Light Blue | Dark Blue | | Dark Blue | Light Blue | Dark Blue | | | Dark Blue | | |
| Reduced post-harvest losses | Dark Blue | Dark Blue | Dark Blue | | | Dark Blue | Dark Blue | Dark Blue | Light Blue | Dark Blue | | Dark Blue | Dark Blue | | Dark Blue | | Light Blue |
| Reduced food waste (consumer or retailer) | Light Blue | Light Blue | Light Blue | | Light Blue | Dark Blue | Dark Blue | Dark Blue | Light Blue | | Dark Blue | Dark Blue | Dark Blue | | Dark Blue | | Light Blue |
| Material substitution | | | | | | Light Blue | Dark Blue | | Light Blue | | Light Blue | Dark Blue | Light Blue | | Light Blue | | |

⁹ FOOTNOTE: Note that this refers to large areas of bioenergy crops capable of producing large mitigation benefits (> 3 GtCO₂ yr⁻¹). The effect of bioenergy and BECCS on SDG is scale and context dependent (see Cross-Chapter Box 7 in this chapter; Section 6.2).

| | | | | | | | | | | | | | | | | | |
|-------------------------------------|------------|------------|------------|------------|------------|------------|------------|-----------|-----------|------------|------------|------------|------------|-------|------------|-----------|------------|
| Sustainable sourcing | Dark Teal | Light Teal | Light Teal | White | White | Light Teal | White | Dark Teal | Dark Teal | Light Teal | Light Teal | Light Teal | White | White | Dark Teal | White | Light Teal |
| Management of supply chains | Dark Teal | Dark Teal | Dark Teal | Light Teal | Dark Teal | Light Red | Dark Teal | Dark Teal | Dark Teal | Light Teal | Dark Teal | Light Teal | Dark Teal | White | Light Teal | White | Light Teal |
| Enhanced urban food systems | Light Teal | Dark Teal | Light Teal | Light Teal | Dark Teal | Light Red | Light Teal | Dark Teal | Dark Teal | Dark Teal | Dark Teal | Dark Teal | Dark Teal | White | Light Teal | Dark Teal | White |
| Improved food processing & retail | Light Teal | Light Teal | Light Teal | White | Light Red | Light Teal | Dark Teal | Dark Teal | Dark Teal | White | Light Teal | Light Teal | Light Teal | White | White | White | Light Teal |
| Improved energy use in food systems | White | Light Teal | Light Teal | White | Light Teal | Light Teal | Light Teal | White | White | White | Light Teal | Light Teal | White | White | White | White | White |

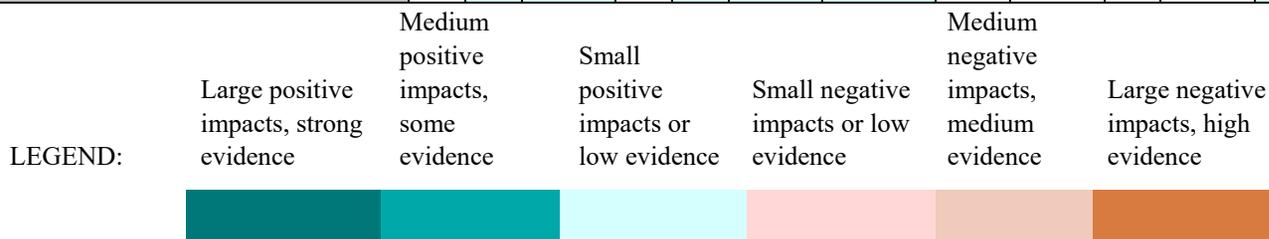


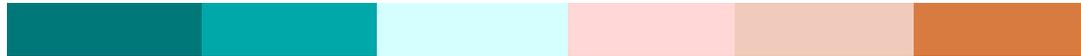
Table 6.75 Impacts on the UN SDG of integrated response options based on risk management

| Integrated response options based on risk management | GOAL 1: No Poverty | GOAL 2: Zero Hunger | GOAL 3: Good Health and Well-being | GOAL 4: Quality Education | GOAL 5: Gender Equality | GOAL 6: Clean Water and Sanitation | GOAL 7: Affordable and Clean Energy | GOAL 8: Decent Work and Economic Growth | GOAL 9: Industry, Innovation and Infrastructure | GOAL 10: Reduced Inequality | GOAL 11: Sustainable Cities and Communities | GOAL 12: Responsible Consumption and Production | GOAL 13: Climate Action | GOAL 14: Life Below Water | GOAL 15: Life on Land | GOAL 16: Peace and Justice Strong Institutions | GOAL 17: Partnerships to achieve the Goal |
|--|--------------------|---------------------|------------------------------------|---------------------------|-------------------------|------------------------------------|-------------------------------------|---|---|-----------------------------|---|---|-------------------------|---------------------------|-----------------------|--|---|
| Management of urban sprawl | Dark Teal | Dark Teal | Dark Teal | White | White | Dark Teal | Dark Teal | Light Red | Dark Teal | Dark Teal | Dark Teal | Dark Teal | Dark Teal | White | Dark Teal | Light Teal | White |
| Livelihood diversification | Dark Teal | Dark Teal | Dark Teal | Dark Teal | Dark Teal | Light Teal | Light Teal | Dark Teal | White | Light Teal | Light Teal | Light Teal | Dark Teal | White | Dark Teal | White | White |
| Use of local seeds | Dark Teal | Light Teal | Dark Teal | White | Dark Teal | Light Teal | White | Dark Teal | White | Light Teal | Light Teal | Dark Teal | Light Teal | White | Dark Teal | Dark Teal | Light Red |
| Disaster risk management | Dark Teal | Dark Teal | Dark Teal | Dark Teal | Dark Teal | Dark Teal | White | Dark Teal | Dark Teal | Light Teal | Dark Teal | Dark Teal | White | Dark Teal | Dark Teal | Dark Teal | White |



LEGEND:

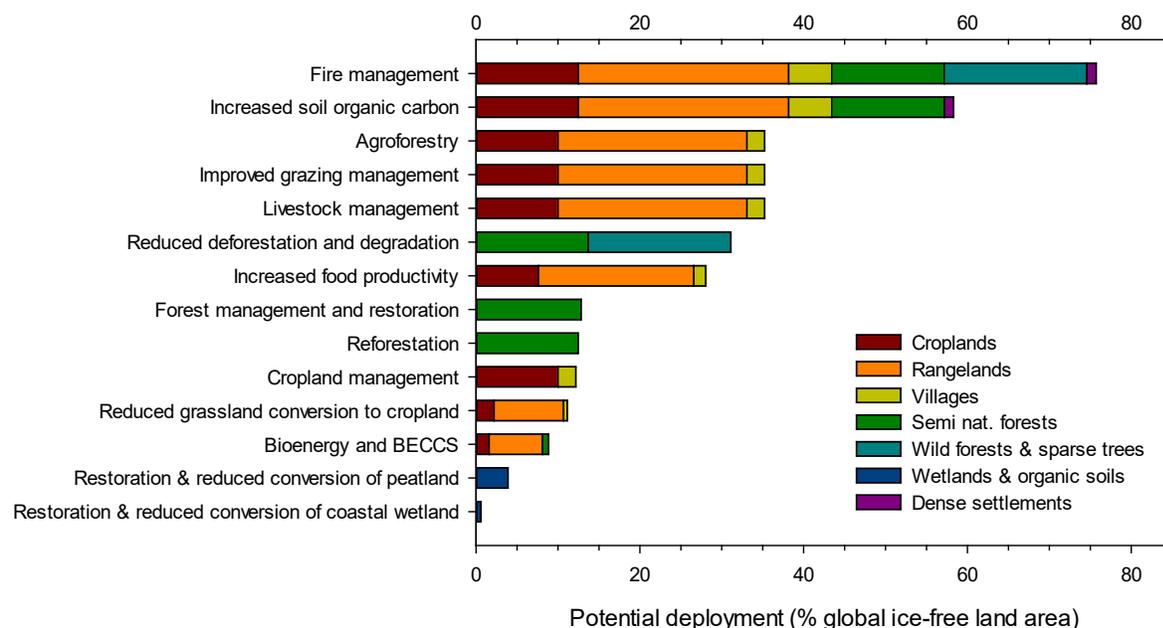
| | | | | | | |
|--|--|--|--|--|--|---------------------------------------|
| | | Medium positive impacts, some evidence | Small positive impacts or low evidence | Small negative impacts or low evidence | Medium negative impacts, medium evidence | Large negative impacts, high evidence |
|--|--|--|--|--|--|---------------------------------------|



1 6.4.4 Opportunities for implementation of Integrated Response Options

2 6.4.4.1 Where can the response options be applied?

3 As shown in Section 6.1.3, a large part of the land area is exposed to overlapping land challenges,
 4 especially in villages, croplands and rangelands. The deployment of land management responses may
 5 vary with local exposure to land challenges. For instance, with croplands exposed to a combination of
 6 land degradation, food insecurity and climate change adaptation challenges, maximising the co-benefits of
 7 land management responses would require selecting responses having only co-benefits for these 3
 8 overlapping challenges, as well as for climate change mitigation which is a global challenge. Based on
 9 these criteria, Figure 6.6 shows the potential deployment area of land management responses across land
 10 use types (or anthromes).



11 Potential deployment (% global ice-free land area)

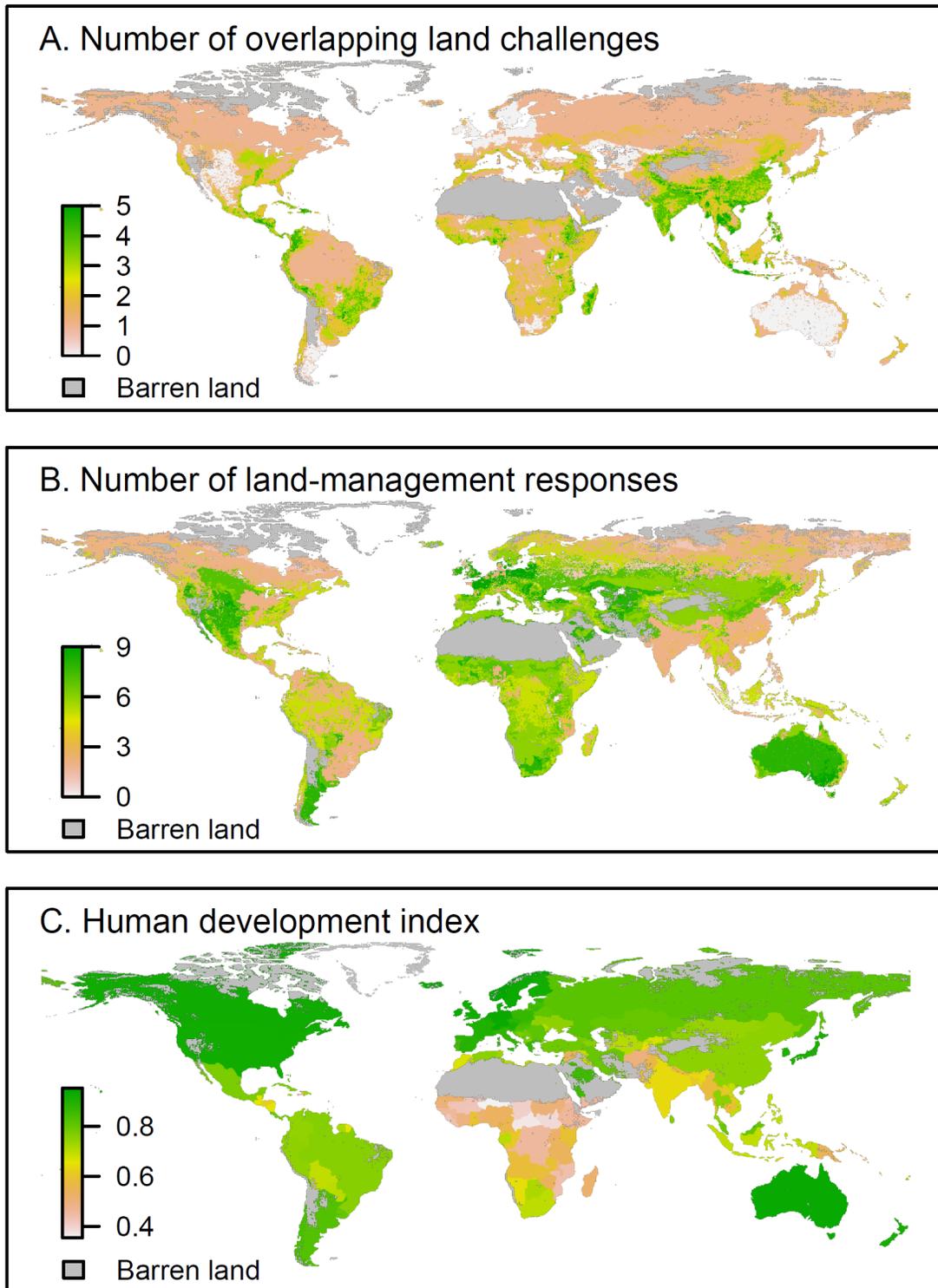
12 **Figure 6.6 Potential deployment area of land management responses (see Table 6.1) across land use types (or**
 13 **anthromes, see section 6.3), when selecting responses having only co-benefits for local challenges and for**
 14 **climate change mitigation and no large adverse side-effect on global food security. See Figure 6.2 for the**
 15 **criteria used to map challenges (desertification, land degradation, climate change adaptation, chronic**
 16 **undernourishment, biodiversity, groundwater stress and water quality) considered. No response option was**
 17 **identified for barren lands.**

18 Land management responses having co-benefits across the range of challenges, including climate change
 19 mitigation, could be deployed between one (coastal wetlands, peatlands, forest management and
 20 restoration, reforestation) and 5 (increased soil organic carbon) or 6 (fire management) land use types
 21 (Figure 6.6). Fire management and increased soil organic carbon have a large potential since they could
 22 be deployed with mostly co-benefits and few adverse effects over 76 and 58% of the ice-free land area. In
 23 contrast, other responses have a limited area-based potential due to biophysical constraints (e.g., limited
 24 extent of organic soils and of coastal wetlands for conservation and restoration responses), or due to the
 25 occurrence of adverse effects. Despite strong co-benefits for climate change mitigation, the deployment of
 26 bioenergy and BECCS would have co-benefits on only 9% of the ice-free land area (Figure 6.6), given
 27 adverse effects of this response option for food security, land degradation, climate change adaptation and
 28 desertification (see Tables 6.62-6.69).

1 Without including the global climate change mitigation challenge, there are up to 5 overlapping
2 challenges on lands which are not barren (Fig. 6.7A, calculated from the overlay of individual challenges
3 shown in Fig. 6.2) and up to 9 land management response options having only co-benefits for these
4 challenges and for climate change mitigation (Fig. 6.7B). Across countries, the mean number of land
5 management response options with mostly co-benefits declines ($p < 0.001$, Spearman rank order
6 correlation) with the mean number of land challenges. Hence, the higher the number of land challenges
7 per country, the fewer the land management response options having only co-benefits for the challenges
8 encountered.

9
10 Enabling conditions (see Section 6.1.2.2) for the implementation of land management responses partly
11 depend upon human development (economics, health and education) as estimated by a country scale
12 composite index, the Human Development index (HDI, United Nations Development Program, 2018)
13 (Figure 6.7C). Across countries, HDI is negatively correlated ($p < 0.001$, Spearman rank order correlation)
14 with the mean number of land challenges. Therefore, on a global average, the higher the number of local
15 challenges faced, the fewer the land management responses having only co-benefits and the lower the
16 human development (Figure 6.7) that could favour the implementation of these responses.

17



1
2
3
4
5
6
7
Figure 6.7 Global distributions of (A) number of overlapping land challenges (desertification, land degradation, climate change adaptation, chronic undernourishment, biodiversity, groundwater stress and water quality, see Fig. 6.2); (B) number of land management responses providing medium to large co-benefits and no adverse side-effects (see Fig. 6.6) across challenges; (C) Human Development Index (HDI) by country.

The Human Development Index (United Nations Development Programme, 2018) is a country based composite statistical index measuring average achievement in three basic dimensions of human development

1 **a long and healthy life (estimated from life expectancy at birth), knowledge (estimated from years of**
 2 **schooling) and a decent standard of living (estimated from gross national income per capita)**

3 **6.4.4.2 Interlinkages and response options in future scenarios**

4 This section assesses more than eighty articles quantifying the effect of various response options in the
 5 future, covering a variety of response options and land-based challenges. These studies cover spatial
 6 scales ranging from global (Popp et al. 2017; Fujimori et al. 2018a) to regional (Calvin et al. 2016a; Frank
 7 et al. 2015) to country-level (Gao and Bryan 2017; Pedercini et al. 2018). This section focuses on models
 8 that can quantify interlinkages between response options, including agricultural economic models, land
 9 system models, and integrated assessment models. The IAM and non-IAM literature, however, is also
 10 categorised separately to elucidate what is and is not included in global mitigation scenarios, like those
 11 included in the SR15. Results from bottom-up studies and models (e.g., Griscom et al. 2017a) are
 12 assessed in Section 6.2-6.3.

13 *Response options in future scenarios:*

14 More than half of the 40 land-based response options discussed in this chapter are represented in global
 15 IAMs models used to develop and analyse future scenarios, either implicitly or explicitly (Table 6.76).
 16 For example, all IAMs include improved cropland management, either explicitly through technologies
 17 that improve N use efficiency (Humpenöder et al. 2018a) or implicitly through marginal abatement cost
 18 curves that link reductions in N₂O emissions from crop production to carbon prices (most other models).

19 However, the literature discussing the effect of these response options on land-based challenges is more
 20 limited (Table 6.76). Fifty-seven studies (forty-three IAM studies) articulate the effect of response options
 21 on mitigation, with most including bioenergy and BECCS or a combination of reduced deforestation,
 22 reforestation, and afforestation. Thirty-seven studies (twenty-one IAM studies) discuss the implications of
 23 response options on food security, usually using food price as a metric. While a small number of non-
 24 IAM studies examine the effects of response options on desertification (three studies) and land
 25 degradation (five studies), no IAM studies were identified. However, some studies quantify these
 26 challenges indirectly using IAMs either via climate outputs from the RCPs (Huang et al. 2016) or by
 27 linking IAMs to other land and ecosystem models (Brink et al. 2018; UNCCD 2017).

28 For many of the scenarios in the literature, land-based response options are included as part of a suite of
 29 mitigation options (Popp et al. 2017; van Vuuren et al. 2015a). As a result, it is difficult to isolate the
 30 effect of an individual option on land-related challenges. A few studies focus on specific response options
 31 (Calvin et al. 2014a; Popp et al. 2014b; Kreidenweis et al. 2016b; Humpenöder et al. 2018a), quantifying
 32 the effect of including an individual option on a variety of sustainability targets.

33
 34 **Table 6.76 Number of IAM and non-IAM Studies Including Specific Response Options (rows) and**
 35 **Quantifying Particular Land Challenges (columns). The third column shows how many IAM models include**
 36 **the individual response option; red indicates all models include the option, orange indicates more than half of**
 37 **all models, yellow indicates less than half, and white indicates no models. The remaining columns show**
 38 **challenges related to climate change (C), mitigation (M), adaptation (A), desertification (D), land degradation**
 39 **(L), food security (F), and biodiversity/ecosystem services/sustainable development (B). The colour indicates**
 40 **the number of total studies, with 0 (white), 1-5 (green), 6-10 (light blue), 11-15 (dark blue), and 16 or more**
 41 **(purple). Additionally, counts of total (left value) and IAM-only (right value) studies are included. Some**
 42 **IAMs include agricultural economic models which can also be run separately; these models are not counted**
 43 **as IAM literature when used on their own. Studies using a combination of IAMs and non-IAMs are included**
 44 **in the total only. A complete list of studies is included in the supplementary material.**

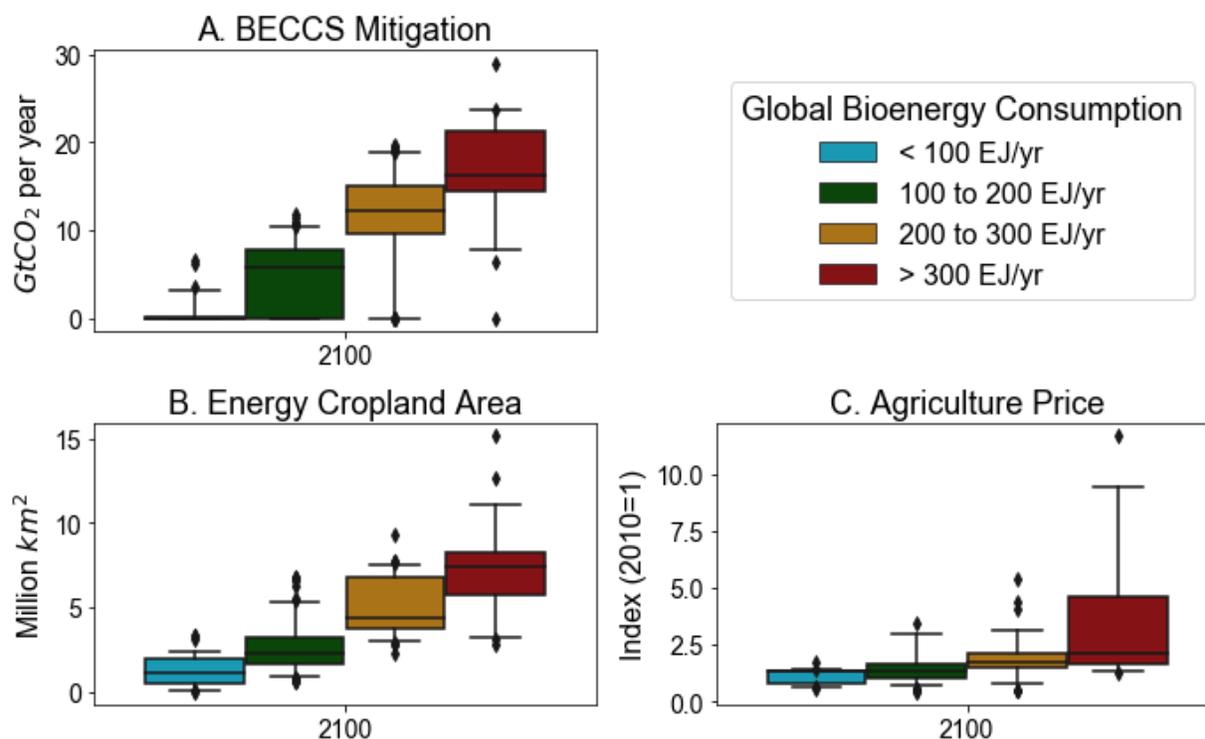
| Category | Response Option | IAMs ^a | Studies [Total/IAM] | | | | | | |
|---------------------------------|--|-------------------|------------------------|-------|-----|-----|----------------|----------------|-------|
| | | | C | M | A | D | L ^b | F ^c | B |
| Land Management | Increased food productivity | | 1/1 | 18/14 | 5/1 | 2/0 | 3/0 | 18/9 | 12/6 |
| | Improved cropland management | | 0/0 | 15/11 | 7/2 | 0/0 | 0/0 | 13/6 | 7/4 |
| | Improved grazing land management | | 0/0 | 1/0 | 1/0 | 0/0 | 0/0 | 1/0 | 0/0 |
| | Improved livestock management | | 0/0 | 10/6 | 1/0 | 2/0 | 2/0 | 7/3 | 5/2 |
| | Agroforestry | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Agricultural diversification | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Reduced grassland conversion to cropland | | 0/0 | 2/2 | 0/0 | 0/0 | 0/0 | 1/1 | 1/1 |
| | Integrated water management | | 1/0 | 17/12 | 5/2 | 0/0 | 2/0 | 13/7 | 20/13 |
| | Improved forest management | | 0/0 | 2/0 | 0/0 | 1/0 | 1/0 | 2/0 | 2/0 |
| | Reduced deforestation and degradation | | 2/2 | 24/20 | 1/0 | 1/0 | 1/0 | 14/9 | 14/8 |
| | Reforestation and forest restoration | | 3/3 | 19/18 | 1/1 | 1/0 | 2/0 | 9/8 | 9/6 |
| | Afforestation | | 3/3 | 24/21 | 2/1 | 0/0 | 0/0 | 10/9 | 8/7 |
| | Increased soil organic carbon content | | 0/0 | 3/1 | 0/0 | 0/0 | 0/0 | 1/1 | 0/0 |
| | Reduced soil erosion | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Reduced soil salinisation | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Reduced soil compaction | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Biochar addition to soil | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Fire management | | 0/0 | 1/1 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Reduced landslides and natural hazards | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Reduced pollution including acidification | | 2/2 | 18/16 | 2/1 | 0/0 | 0/0 | 10/7 | 6/6 |
| | Management of invasive species / encroachment | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| | Restoration and reduced conversion of coastal wetlands | | 0/0 | 0/0 | 0/0 | 1/0 | 1/0 | 0/0 | 1/0 |
| | Restoration and reduced conversion of peatlands | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
| Biodiversity conservation | | 1/0 | 7/3 | 0/0 | 1/0 | 3/0 | 4/2 | 8/1 | |
| Enhanced weathering of minerals | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| Bioenergy and BECCS | | 5/4 | 50/40 | 7/4 | 0/0 | 2/0 | 25/18 | 21/13 | |
| Value Chain Management | Dietary change | | 0/0 | 15/12 | 1/0 | 2/0 | 2/0 | 13/9 | 10/7 |
| | Reduced post-harvest losses | | 0/0 | 5/4 | 0/0 | 0/0 | 2/2 | 2/1 | |
| | Reduced food waste (consumer or retailer) | | 0/0 | 6/4 | 0/0 | 0/0 | 4/2 | 3/1 | |
| | Material substitution | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| | Sustainable sourcing | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| | Management of supply chains | | 1/1 | 11/9 | 8/1 | 2/0 | 3/0 | 17/9 | 7/3 |
| | Enhanced urban food systems | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| | Improved food processing and retailing | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| Risk Management | Improved energy use in food systems | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| | Management of urban sprawl | | 0/0 | 0/0 | 0/0 | 1/0 | 1/0 | 0/0 | 1/0 |
| | Livelihood diversification | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| | Use of local seeds | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |
| | Disaster risk management | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | |

| | | | | | | | | | |
|--|--------------------------|--|-----|-----|-----|-----|-----|-----|-----|
| | Risk sharing instruments | | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 | 0/0 |
|--|--------------------------|--|-----|-----|-----|-----|-----|-----|-----|

1 ^a Only IAMs that are used in the papers assessed are included in this column.
 2 ^b There are many indicators for land degradation (see Chapter 4). In this table, studies are categorised as quantifying
 3 land degradation if they explicitly discuss land degradation.
 4 ^c Studies are categorised as quantifying food security if they report food prices or the population at risk of hunger.
 5

6 *Interactions and Interlinkages between Response Options:*

7 The effect of response options on desertification, land degradation, food security, biodiversity, and other
 8 sustainable development goals depends strongly on which options are included and the extent to which
 9 they are deployed. For example, sections 2.6, 6.3.6, and the Cross-Chapter Box 7 on Bioenergy note that
 10 bioenergy and BECCS has a large mitigation potential but could potentially have adverse side effects for
 11 land degradation, food security, and other sustainable development goals. Global modelling studies
 12 demonstrate that these effects are dependent on scale. Increased use of bioenergy can result in increased
 13 mitigation (Figure 6.8, Panel A) and reduced climate change, but can also lead to increased energy
 14 cropland expansion (Figure 6.8, Panel B), and increased competition for land resulting in increased food
 15 prices (Figure 6.8, Panel C). However, the exact relationship between bioenergy deployment and each
 16 sustainability target depends a number of other factors, including the feedstock used, the underlying
 17 socioeconomic scenario, assumptions about technology and resource base, the inclusion of other response
 18 options, and the specific model used (Calvin et al. 2014a; Clarke and Jiang 2014b; Popp et al. 2014b,
 19 2017; Kriegler et al. 2014).



20
 21 **Figure 6.8 Correlation between Bioenergy Use and Other Indicators.** Panel A shows global CO₂ sequestration
 22 by BECCS in 2100. Panel B shows global energy cropland area in 2100. Panel C shows agricultural prices in
 23 2100 indexed to 2010. Data are binned based on the amount of bioenergy used globally in 2100. All scenario
 24 data that include both bioenergy consumption and the variable of interest are included in the figure; the
 25 resulting number of scenarios varies per panel with 352 in panel A, 262 in panel B, and 172 in panel C. The

1 **boxes represent the interquartile range (i.e., the middle 50% of all scenarios), the line in the middle of the box**
2 **represents the median, and the whiskers represent the 5 to 95% range of scenarios. Data is from an update of**
3 **the IAMC Scenario Explorer developed for the SR15 (Huppmann et al. 2018; Rogelj et al. 2018b).**

4 The previous sections have examined the effects of individual land-response options on multiple
5 challenges. A number of studies using global modelling and analyses have examined interlinkages and
6 interaction effects among land response options by incrementally adding or isolating the effects of
7 individual options. Most of these studies focus on interactions with bioenergy and BECCS (Table 6.77).
8 Adding response options that require land (e.g., reforestation, afforestation, reduced deforestation,
9 avoided grassland conversion, or biodiversity conservation), results in increased food prices (Calvin et al.
10 2014a; Humpenöder et al. 2014a; Obersteiner et al. 2016a; Reilly et al. 2012a) and potentially increased
11 temperature through biophysical climate effects (Jones et al. 2013). However, this combination can result
12 in reduced water consumption (Hejazi et al. 2014c), reduced cropland expansion (Calvin et al. 2014a;
13 Humpenöder et al. 2018a), increased forest cover (Calvin et al. 2014a; Humpenöder et al. 2018a; Wise et
14 al. 2009a) and reduced biodiversity loss (Pereira et al. 2010), compared to scenarios with bioenergy and
15 BECCS alone. While these options increase total mitigation, they reduce mitigation from bioenergy and
16 BECCS as they compete for the same land (Wu et al. 2019; Baker et al. 2019a; Calvin et al. 2014a;
17 Humpenöder et al. 2014a).

18 The inclusion of land-sparing options (e.g., dietary change, increased food productivity, reduced food
19 waste, management of supply chains) in addition to bioenergy and BECCS results in reduced food prices,
20 reduced agricultural land expansion, reduced deforestation, reduced mitigation costs, reduced water use,
21 and reduced biodiversity loss (Bertram et al. 2018; Wu et al. 2019; Obersteiner et al. 2016a; Stehfest et al.
22 2009; van Vuuren et al. 2018a). These options can increase bioenergy potential, resulting in increased
23 mitigation than from bioenergy and BECCS alone (Wu et al. 2019; Stehfest et al. 2009; Favero and
24 Massetti 2014).

25 Other combinations of land response options create synergies, alleviating land pressures. The inclusion of
26 increased food productivity and dietary change can increase mitigation, reduce cropland use, reduce water
27 consumption, reduce fertiliser application, and reduce biodiversity loss (Springmann et al. 2018c;
28 Obersteiner et al. 2016a). Similarly, improved livestock management combined with increased food
29 productivity can reduce agricultural land expansion (Weindl et al. 2017). Reducing disturbances (e.g., fire
30 management) in combination with afforestation can increase the terrestrial carbon sink, resulting in
31 increased mitigation potential and reduced mitigation cost (Le Page et al. 2013a).

32 Studies including multiple land response options often find that the combined mitigation potential is not
33 equal to the sum of individual mitigation potential as these options often share the same land. For
34 example, including both afforestation and bioenergy and BECCS results in a cumulative reduction in
35 GHG emissions of 1200 GtCO₂ between 2005 and 2100, which is much lower than the sum of the
36 contributions of bioenergy (800 GtCO₂) and afforestation (900 GtCO₂) individually (Humpenöder et al.
37 2014a). More specifically, Baker et al. (2019a) find that woody bioenergy and afforestation are
38 complementary in the near-term, but become substitutes in the long-term, as they begin to compete for the
39 same land. Similarly, the combined effect of increased food productivity, dietary change, and reduced
40 waste on GHG emissions is less than the sum of the individual effects (Springmann et al. 2018c).

1 **Table 6.77 Interlinkages between bioenergy and BECCS and other response options.** Table indicates the
 2 **combined effects of multiple land-response options on climate change (C), mitigation (M), adaptation (A),**
 3 **desertification (D), land degradation (L), food security (F), and biodiversity/ecosystem services/sustainable**
 4 **development (O).** Each cell indicates the implications of adding the option specified in the row in addition to
 5 **bioenergy and BECCS.** Blue colours indicate positive interactions (e.g., including the option in the second
 6 **column increases mitigation, reduces cropland area, or reduces food prices relative to bioenergy and BECCS**
 7 **alone).** Red colours indicate negative interactions; yellow indicates mixed interactions (some positive, some
 8 **negative).** Note that only response option combinations found in the assessed literature are included in the
 9 **interest of space.**

| | C ^a | M ^b | A | D | L ^c | F | O ^d | Context and Sources |
|--|----------------|----------------|---|---|----------------|------|----------------|--|
| Increased food productivity | | Blue | | | Blue | Blue | Blue | Sources: Humpenöder et al. 2018a; Obersteiner et al. 2016a |
| Increased food productivity; improved livestock management | | | | | Blue | | | Sources: van Vuuren et al. 2018a |
| Improved cropland management | | | | | | | Blue | Sources: Humpenöder et al. 2018a |
| Integrated water management | | Red | | | Red | Red | Yellow | O: Reduces water use, but increases fertiliser use. Sources: Humpenöder et al. 2018a |
| Reduced deforestation | | Blue | | | Blue | | | Sources: Calvin et al. 2014a; Humpenöder et al. 2018a |
| Reduced deforestation, Avoided grassland conversion | | Blue | | | Blue | Red | Yellow | O: Reduces biodiversity loss and fertiliser, but increases water use. Sources: Calvin et al. 2014a; Obersteiner et al. 2016a |
| Reforestation | | | | | | Red | | Sources: Reilly et al. 2012a |
| Reforestation, Afforestation, Avoided grassland conversion | Red | Blue | | | Blue | Red | Blue | Sources: Calvin et al. 2014a; Hejazi et al. 2014a; Jones et al. 2013 |
| Afforestation | | Blue | | | | | | Sources: Humpenöder et al. 2014a |
| Biodiversity conservation | | Yellow | | | Blue | Red | Yellow | M: Reduces emissions but also reduces bioenergy potential. O: Reduces biodiversity loss but increases water use. Sources: Obersteiner et al. 2016a; Wu et al. 2019 |
| Reduced pollution | | | | | Blue | | | Sources: van Vuuren et al. 2018a |
| Dietary change | | Blue | | | | Blue | | Sources: Bertram et al. 2018; Stehfest et al. 2009; Wu et al. 2019 |
| Reduced food waste; dietary change | | | | | Blue | | | Sources: van Vuuren et al. 2018a |
| Management of supply chains | | Blue | | | | | | Sources: Favero and Massetti 2014 |
| Management of supply chains; increased productivity | | Blue | | | | | | Sources: Wu et al. 2019 |
| Reduced deforestation; Improved cropland management; Improved food productivity; Integrated water management | | Blue | | | Blue | Blue | Blue | Sources: Humpenöder et al. 2018a |
| Reduced deforestation; Management of Supply Chains; Integrated Water Management; Improved cropland management; Increased food productivity | | | | | Blue | Blue | Blue | Sources: Bertram et al. 2018 |
| Reduced deforestation; Management of Supply Chains; Integrated Water Management; Improved cropland management; | | | | | Blue | Blue | Blue | Sources: Bertram et al. 2018 |

| | | | | | | | |
|--|--|--|--|--|--|--|--|
| Increased food productivity; dietary change | | | | | | | |
|--|--|--|--|--|--|--|--|

^a Includes changes in biophysical effects on climate (e.g., albedo)

^b Either through reduced emissions, increased mitigation, reduced mitigation cost, or increased bioenergy potential. For increased mitigation, a positive indicator in this column only indicates that total mitigation increases and not that the total is greater than the sum of the individual options.

^c Uses changes in cropland or forest as an indicator (reduced cropland expansion or reduced deforestation are considered positive)

^d Includes changes in water use or scarcity, fertiliser use, or biodiversity

Land-related response options can also interact with response options in other sectors. For example, limiting deployment of a mitigation response option will either result in increased climate change or additional mitigation in other sectors. A number of studies have examined limiting bioenergy and BECCS. Some such studies show increased emissions (Reilly et al. 2012a). Other studies meet the same climate goal, but reduce emissions elsewhere *via* reduced energy demand (Grubler et al. 2018; van Vuuren et al. 2018a), increased fossil CCS, nuclear energy, energy efficiency and/or renewable energy (van Vuuren et al. 2018a; Rose et al. 2014b; Calvin et al. 2014a; van Vuuren et al. 2017b), dietary change (van Vuuren et al. 2018a), reduced non-CO₂ emissions (van Vuuren et al. 2018a), or lower population (van Vuuren et al. 2018a). The co-benefits and adverse side-effects of non-land mitigation options are discussed in SR15, Chapter 5. Limitations on bioenergy and BECCS can result in increases in the cost of mitigation (Kriegler et al. 2014; Edmonds et al. 2013a). Studies have also examined limiting CDR, including reforestation, afforestation, and bioenergy and BECCS (Kriegler et al. 2018a,b). These studies find that limiting CDR can increase mitigation costs, increase food prices, and even preclude limiting warming to less than 1.5°C above pre-industrial levels (Kriegler et al. 2018a,b; Muratori et al. 2016).

In some cases, the land challenges themselves may interact with land-response options. For example, climate change could affect the production of bioenergy and BECCS. A few studies examine these effects, quantifying differences in bioenergy production (Calvin et al. 2013a; Kyle et al. 2014) or carbon price (Calvin et al. 2013a) as a result of climate change. Kyle et al. (2014) finds increase in bioenergy production due to increases in bioenergy yields, while Calvin et al. (2013a) finds declines in bioenergy production and increases in carbon price due to the negative effects of climate on crop yield.

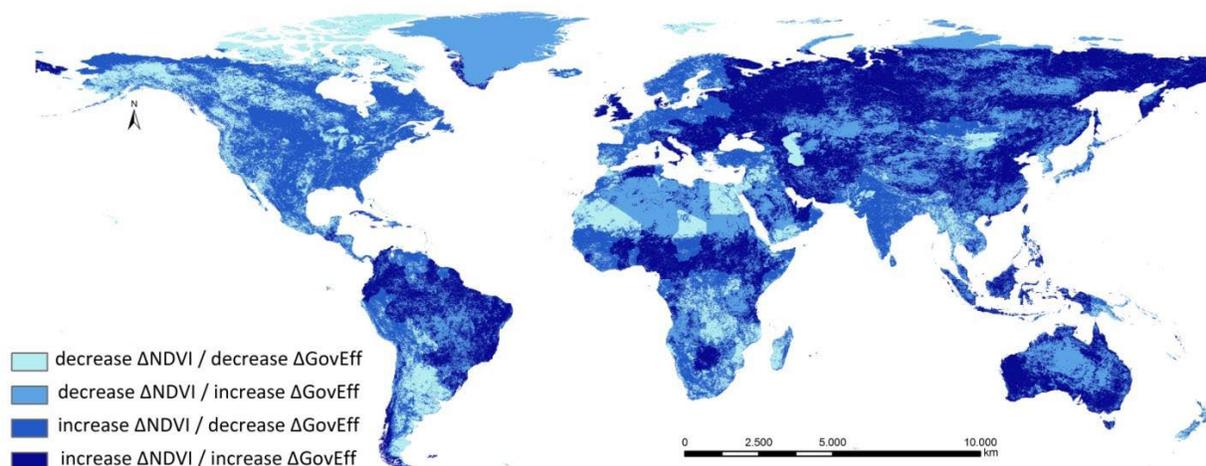
Gaps in the Literature:

Not all of the response options discussed in this chapter are included in the assessed literature, and many response options are excluded from the IAM models. The included options (e.g. bioenergy and BECCS; reforestation) are some of the largest in terms of mitigation potential (see Section 6.3). However, some of the options excluded also have large mitigation potential. For example, biochar, agroforestry, restoration/avoided conversion of coastal wetlands, and restoration/avoided conversion of peatland all have mitigation potential of ~1 GtCO₂ yr⁻¹ (Griscom et al. 2017). Additionally, quantifications of and response options targeting land degradation and desertification are largely excluded from the modelled studies, with a few notable exceptions (Wolff et al. 2018; Gao and Bryan 2017; Brink et al. 2018; UNCCD 2017). Finally, while a large number of papers have examined interactions between bioenergy and BECCS and other response options, the literature examining other combinations of response options is more limited.

6.4.4.3 Resolving challenges in response option implementation

The 40 response options assessed in this chapter face a variety of barriers to implementation that require action across multiple actors to overcome (section 6.4.1). Studies have noted that while adoption of

1 response options by individuals may depend on individual assets and motivation, larger structural and
 2 institutional factors are almost always equally important if not more so (Adimassu et al. 2016; Djenontin
 3 et al. 2018), though harder to capture in research variables (Schwilch et al. 2014). These institutional and
 4 governance factors can create an enabling environment for SLM practices, or challenges to their
 5 adoption (Adimassu et al. 2013). Governance factors include the institutions that manage rules and
 6 policies, the social norms and collective actions of participants (including civil society actors and the
 7 private sector), and the interactions between them (Ostrom 1990; Huntjens et al. 2012; Davies 2016).
 8 Many of Ostrom’s design principles for successful governance can be applied to response options for
 9 SLM; these principles are: (1) clearly defined boundaries; (2) understanding of both benefits and costs;
 10 (3) collective choice arrangements; (4) monitoring; (5) graduated sanctions; (6) conflict-resolution
 11 mechanisms; (7) recognition of rights; and (8) nested (multi-scale) approaches. Unfortunately, studies of
 12 many natural resources and land management policy systems in developing countries in particular often
 13 show the opposite: a lack of flexibility, strong hierarchical tendencies, and a lack of local participation in
 14 institutional frameworks (Ampaire et al. 2017). Analysis of government effectiveness (GE)– defined as
 15 quality of public services, policy formulation and implementation, civil service and the degree of its
 16 independence from political pressures as well as credibility of the government’s commitment to its
 17 policies (Kaufman et al. 2010) – has been shown to play a key role in land management. GE mediates
 18 land user actions on land management and investment, and government policies and laws can help land
 19 users adopt sustainable land management practices (Nkonya et al. 2016) (Figure 6.9).



20
 21 **Figure 6.9 Relationship between changes in government effectiveness and changes in land management**

22 Notes: Δ NDVI = Change in Normalized Difference Vegetation Index (baseline year 2001, Endline year 2010).
 23 Source of NDVI data: MODIS Δ GovEff = Change in Government effectiveness (baseline year 2001, Endline year
 24 2010). Source of Government effectiveness: World Bank. Source: Nkonya et al 2016.

25 It is simply not a matter of putting the ‘right’ institutions or policies in place, however, as governance can
 26 be undermined by inattention to power dynamics (Fabinyi et al. 2014). Power shapes how actors gain
 27 access and control over resources, and negotiate, transform and adopt certain response options or not.
 28 These variable dynamics of power between different levels and stakeholders have an impact on the ability
 29 to implement different response options. The inability of many national governments to address social
 30 exclusion in general will have an effect on the implementation of many response options. Further,
 31 response options themselves can become avenues for actors to exert power claims over others

1 (Nightingale 2017). For example, there have been many concerns that reduced deforestation and
2 degradation projects run the risk of reversing trends towards decentralisation in forest management and
3 create new power disparities between the state and local actors (Phelps et al. 2010). Below we assess how
4 two important factors, the involvement of stakeholders and the coordination of action across scales, will
5 help in moving from response options to policy implementation, a theme chapter 7 takes up in further
6 detail.

7 *Involvement of stakeholders*

8 There are a wide range of stakeholders that are necessary for successful land, agricultural and
9 environmental policy, and implementing response options requires that a range of actors, including
10 businesses, consumers, land managers, indigenous and local communities, scientists, and policymakers
11 work together for success. Diverse stakeholders have a particularly important role to play in defining
12 problems, assessing knowledge and proposing solutions (Phillipson et al. 2012; Stokes et al. 2006). Lack
13 of connection between science knowledge and on-the-ground practice has hampered adoption of many
14 response options in the past; simply presenting ‘scientifically’ derived response options is not enough
15 (Marques et al. 2016). For example, the importance of recognising and incorporating local knowledge
16 (LK) and indigenous knowledge (IK) is increasingly emphasised in successful policy implementation (see
17 Cross-Chapter Box 13 on Indigenous Knowledge, Chapter 7), as local practices of water management,
18 soil fertility management, improved grazing, restoration and sustainable management of forests are often
19 well-aligned with response options assessed by scientists (Marques et al. 2016).

20 Stakeholder engagement is an important approach for successful environmental and climate policy and
21 planning. Tools such as stakeholder mapping, in which affected and interested parties are identified and
22 described in terms of their interrelationships and current or future objectives and aspirations, and
23 scenario-based stakeholder engagement, which combines stakeholder analysis with climate scenarios, are
24 increasingly being applied to facilitate better planning outcomes (Tompkins et al. 2008; Pomeroy &
25 Douvere 2008; Star et al. 2016). Facilitated dialogues early in design processes have shown good success
26 in bringing multiple and sometimes conflicting stakeholders to the table to discuss synergies and trade-
27 offs around policy implementation (Gopnik et al 2012). Knowledge exchange, social learning, and other
28 concepts are also increasingly being incorporated into understandings of how to facilitate sustainable land
29 management (Djenontin et al. 2018), as evidence suggests that negotiating the complexity of SESs
30 requires flexible learning arrangements in particular for multiple stakeholders (Gerlak and Heikkila 2011;
31 Armitage et al. 2018; Heikkila and Gerlak 2018). Social learning has been defined as “a change in
32 understanding and skills that becomes situated in groups of actors/communities of practice through social
33 interactions,” (Albert et al. 2012), and social learning is often linked with attempts to increase levels of
34 participation in decision making, from consultation to more serious community control (Collins and Ison
35 2009; McCrum et al. 2009). Learning also facilitates responses to emerging problems and helps actors in
36 SESs grapple with complexity. One outcome of learning can be adaptive risk management (ARM), in
37 which “one takes action based on available information, monitors what happens, learns from the
38 experience and adjusts future actions based on what has been learnt” (Bidwell et al. 2013). Suggestions to
39 facilitate social learning, ARM, and decision-making include extending science-policy networks and
40 using local bridging organisations, such as extension services, for knowledge co-production (Bidwell et
41 al. 2013; Böcher and Krott 2014; Howarth and Monasterolo 2017) see further discussion in Chapter 7,
42 section 7.5 on Decision-making for Climate and Land).

43

1 Insuring that women are included as key stakeholders in response option implementation is also
2 important, as gender norms and roles affect vulnerability and access to resources, and gender inequality
3 limits the possible range of responses for adoption by women (Lambrou and Piana 2006). For example,
4 environmental change may increase women’s workload as their access to natural resources may decline,
5 or they may have to take up low-wage labour if agriculture becomes unsuitable in their local areas under
6 climate change (Nelson et al. 2002). Every response option considered in this chapter potentially has a
7 gender dimension to it that needs to be taken into consideration (Tables 6.73–6.75 note how response
8 options intersect with SDG 5 Gender Equity); for example, to address food security through sustainable
9 intensification will clearly have to address women farmers in Africa (Kondylis et al. 2016; Garcia and
10 Wanner 2017) (For further information, see Cross-Chapter Box 11: Gender, in Chapter 7).

11 *Challenges of coordination*

12 Coordinated action to implement the response options will be required across a range of actors, including
13 business, consumers, land managers, indigenous and local communities and policymakers to create
14 enabling conditions. Conjoining response options to maximise social, climatic and environmental
15 benefits will require framings of such actions as strong pathways to sustainable development (Ayers and
16 Dodman 2010). As the chapter has pointed out, there are many potentials for synergies, especially among
17 several response options that might be applied together and in coordination with one another (such as
18 dietary change and improved land management measures). This coordination will help ensure that
19 synergies are met and trade-offs minimised, but this will require deliberate coordination across multiple
20 scales, actors and sectors. For example, there are a variety of response options available at different
21 scales that could form portfolios of measures applied by different stakeholders from farm to international
22 scales. Agricultural diversification and use of local seeds by smallholders can be particularly useful
23 poverty reduction and biodiversity conservation measures, but are only successful when higher scales,
24 such as national and international markets and supply-chains, also value these goods in trade regimes,
25 and consumers see the benefits of purchasing these goods. However, the land and food sectors face
26 particular challenges of institutional fragmentation, and often suffer from a lack of engagement between
27 stakeholders at different scales (Biermann et al. 2009; Deininger et al 2014) (see section 7.6.2, Chapter
28 7).

29 Many of the response options listed in this chapter could be potentially implemented as ‘community-
30 based’ actions, including community-based reforestation, community-based insurance, or community-
31 based disaster risk management. Grounding response options in community approaches aims to identify,
32 assist and implement activities “that strengthen the capacity of local people to adapt to living in a riskier
33 and less predictable climate” (Ayers and Forsyth 2009). Research that shows that people willingly come
34 together to provide mutual aid and protection against risk, to manage natural resources, and to work
35 cooperatively to find solutions to environmental provisioning problems. Some activities that fall under
36 this type of collective action can include the creation of institutions or rules; working cooperatively to
37 manage a resource by restricting some activities and encouraging others; sharing information to improve
38 public goods; or mobilising resources, such as capital, to fix a collective problem (Ostrom 2000; Poteete
39 and Ostrom 2004); or engagement in participatory land use planning (Bourgoin 2012; Evers and
40 Hofmeister 2011). These participatory processes “are likely to lead to more beneficial environmental
41 outcomes through better informed, sustainable decisions, and win-win solutions regarding economic and
42 conservation objectives” (Vente et al. 2016), and evaluations of community-based response options have
43 been generally positive (Karim and Thiel 2017a; Tompkins & Adger 2004).

1 Agrawal (2001) has identified more than 30 different indicators that have been important in understanding
2 who undertakes collective action for the environment, including the size of the group undertaking action;
3 the type and distribution of the benefits from the action; the heterogeneity of the group; the dependence of
4 the group on these benefits; the presence of leadership; presence of social capital and trust; and autonomy
5 and independence to make and enforce rules. Alternatively, when households expect the government to
6 undertake response actions, they have less incentive to join in collective action, as the state role has
7 ‘crowded out’ local cooperation (Adger 2009). High levels of social trust and capital can increase
8 willingness of farmers to engage in response options, such as improved soil management or carbon
9 forestry (Stringer et al. 2012; Lee 2017), and social capital helps with connectivity across levels of SESs
10 (Brondizio et al. 2009). (Dietz et al. 2013) lay out important policy directions for more successful
11 facilitation of collective action across scales and stakeholders. These include: providing information;
12 dealing with conflict; inducing rule compliance; providing physical, technical or institutional
13 infrastructure; and being prepared for change. The adoption of participatory protocols and structured
14 processes to select response options together with stakeholders will likely lead to greater success in
15 coordination and participation (Bautista et al. 2017; Franks 2010; Schwilch et al. 2012a).

16 However, wider adoption of community-based approaches is potentially hampered by several factors: the
17 fact that most are small scale (Forsyth 2013; Ensor et al. 2014) and it is often unclear how to assess
18 criteria of success (Forsyth 2013). Others also caution that community-based approaches often are not
19 able to adequately address the key drivers of vulnerability such as inequality and uneven power relations
20 (Nagoda and Nightingale 2017).

21 *Moving from response options to policies*

22 Chapter 7 discusses in further depth the risks and challenges involved in formulating policy responses that
23 meet the demands for sustainable land management and development outcomes, such as food security,
24 community adaptation and poverty alleviation. Chapter 7 in Table 7.1 maps how specific response options
25 might be turned into policies; for example, to implement a response option aimed at agricultural
26 diversification, a range of policies from elimination of agricultural subsidies (which might favour single
27 crops) to environmental farm programs and agro-environmental payments (to encourage alternative
28 crops). Oftentimes, any particular response option might have a variety of potential policy pathways that
29 might address different scales or stakeholders or take on different aspects of coordination and integration
30 (section 7.6.1). Given the unique challenges of decision-making under uncertainty in future climate
31 scenarios, Chapter 7 particularly discusses the need for flexible, iterative, and adaptive processes to turn
32 response options into policy frameworks.

33

34 **Cross-Chapter Box 9: Illustrative Climate and Land Pathways**

Katherine Calvin (The United States of America), Edouard Davin (France/Switzerland), Margot Hurlbert (Canada), Jagdish Krishnaswamy (India), Alexander Popp (Germany), Prajal Pradhan (Nepal/Germany)

Future development of socioeconomic factors and policies influence the evolution of the land-climate system, among others in terms of the land used for agriculture and forestry. Climate mitigation policies can also have a major impact on land use, especially in scenarios consistent with the climate targets of the

Paris Agreement. This includes the use of bio-energy or Carbon Dioxide Removal (CDR), such as bioenergy with carbon dioxide capture and storage (BECCS) and afforestation. Land-based mitigation options have implications for GHG fluxes, desertification, land degradation, food insecurity, ecosystem services and other aspects of sustainable development.

Illustrative Futures

The three illustrative futures are based on the Shared Socioeconomic Pathways (SSPs; (O'Neill et al. 2014c; Riahi et al. 2017b; Popp et al. 2017; Rogelj et al. 2018b); Cross-Chapter Box 1 in Chapter 1). SSP1 is a scenario with a broad focus on sustainability including a focus on human development, technological development, nature conservation, globalised economy, economic convergence and early international cooperation including moderate levels of trade. The scenario assumes a low population growth, relatively high agricultural yields and a move towards less-meat intensive diets (van Vuuren et al. 2017b). Dietary change and reductions in food waste reduce agricultural demands and well-managed land systems enable reforestation and/or afforestation. SSP2 is a scenario in which societal as well as technological development follows historical patterns (Fricko et al. 2017). Land-based CDR is achieved through bioenergy and BECCS, and to a lesser degree by afforestation and reforestation. SSP3 is a scenario with limited technological progress and land-use regulation. Agricultural demands are high due to resource-intensive consumption and a regionalised world leads to reduced flows for agricultural goods. In SSP3, forest mitigation activities and abatement of agricultural GHG emissions are limited due to major implementation barriers such as low institutional capacities in developing countries and delayed as a consequence of low international cooperation (Fujimori et al. 2017a). Emissions reductions are achieved primarily through the energy sector, including the use of bioenergy and BECCS.

Policies in the Illustrative Futures

SSPs are complemented by a set of shared policy assumptions (Kriegler et al. 2014), indicating the types of policies that may be implemented in each future world. IAMs represent the effect of these policies on the economy, energy system, land use and climate with the caveat that they are assumed to be effective or in some cases the policy goals (e.g., dietary change) are imposed rather than explicitly modelled. In the real world, there are various barriers that can make policy implementation more difficult (see 7.4.9). These barriers will be generally higher in SSP3 than SSP1.

SSP1: A number of policies could support this SSP1 future including: effective carbon pricing, emission trading schemes (including net CO₂ emissions from agriculture), carbon taxes, regulations limiting GHG emissions and air pollution, forest conservation (mix of land-sharing and land sparing) through participation, incentives for ecosystem services and secure tenure, and protecting the environment, microfinance, crop and livelihood insurance, agriculture extension services, agricultural production subsidies, low export tax and import tariff rates on agricultural goods, dietary awareness campaigns, regulations to reduce and taxes on food waste, improved shelf life, sugar/fat taxes, and instruments supporting sustainable land management including payment for ecosystem services, land use zoning, REDD+, standards and certification for sustainable biomass production practices, legal reforms on land ownership and access, legal aid, legal education, including reframing these policies as entitlements for women and small agricultural producers (rather than sustainability) (O'Neill et al. 2017; van Vuuren et al. 2017b) (see 7.4).

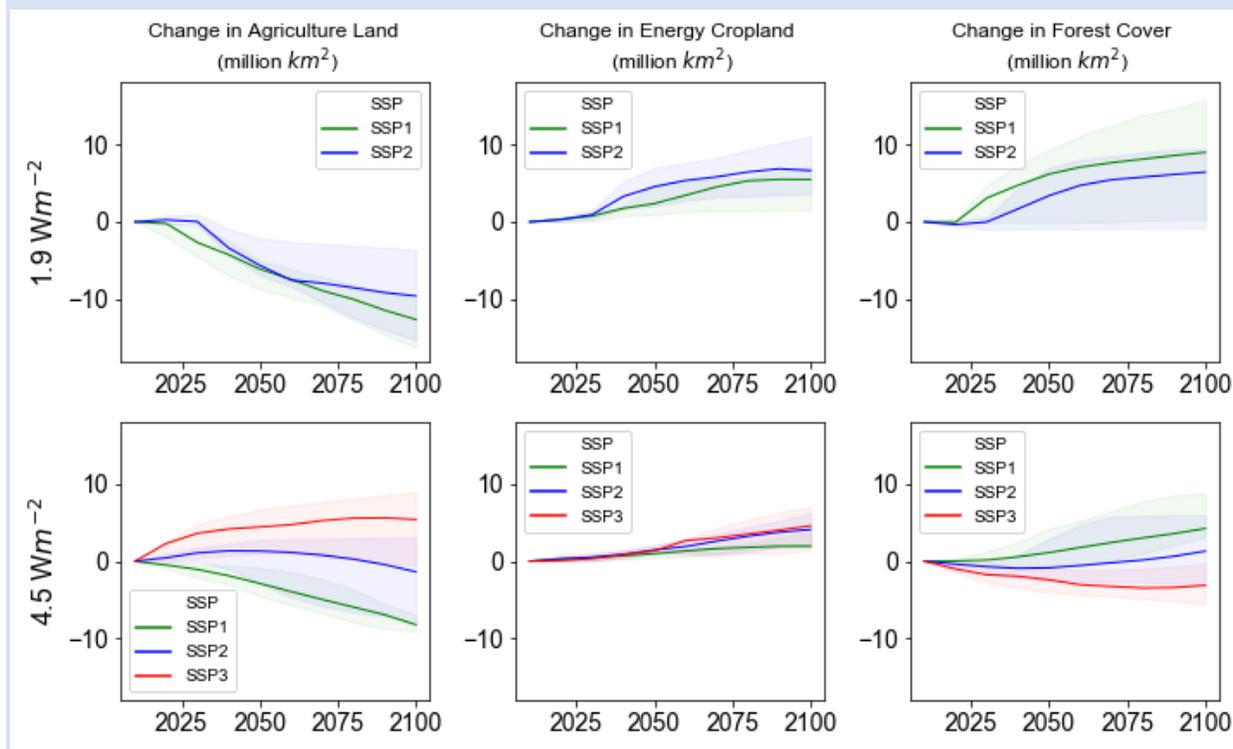
SSP2: The same policies that support the SSP1 could support the SSP2 but may be less effective and only moderately successful. Policies may be challenged by adaptation limits (7.4.9), inconsistency in formal and informal institutions in decision making (7.5.1) or result in maladaptation (7.4.7). Moderately successful sustainable land management policies result in some land competition. Land degradation

neutrality is moderately successful. Successful policies include those supporting bioenergy and BECCS (Rao et al. 2017; Riahi et al. 2017b; Fricko et al. 2017) (see 7.4.6).

SSP3: Policies that exist in SSP1 may or may not exist in SSP3, and are ineffective (O'Neill et al. 2014c). There are challenges to implementing these policies, as in SSP2. In addition, ineffective sustainable land management policies result in competition for land between agriculture and mitigation. Land degradation neutrality is not achieved (Riahi et al. 2017b). Successful policies include those supporting bioenergy and BECCS (see 7.4.6) (Kriegler et al. 2017; Fujimori et al. 2017a; Rao et al. 2017). Demand side food policies are absent and supply side policies predominate. There is no success in advancing land ownership and access policies for agricultural producer livelihood (7.6.5).

Land use and land cover change

Agricultural area in SSP1 declines as a result of the low population growth, agricultural intensification, low meat consumption, and low food waste. In contrast, SSP3 has high population and strongly declining rates of crop yield growth over time, resulting in increased agricultural land area. The SSP2 falls somewhere in between, with its modest growth in all factors. In the climate policy scenarios consistent with the Paris Agreement, bioenergy/BECCS and reforestation/afforestation play an important role in SSP1 and SSP2. The use of these options, and the impact on land, is larger in scenarios that limit radiative forcing in 2100 to 1.9 Wm^{-2} than in the 4.5 Wm^{-2} scenarios. In SSP3, the expansion of land for agricultural production implies that the use of land-related mitigation options is very limited, and the scenario is characterised by continued deforestation.



Cross-Chapter Box 9 Figure 1: Changes in agricultural land (left), energy cropland (middle) and forest cover (right) under three different SSPs (colours) and two different warming levels (rows). Agricultural land includes both pasture and non-energy cropland. Colours indicate SSPs, with SSP1 shown in green, SSP2 in blue, and SSP3 in red. Shaded area show the range across all IAMs; lines show the median across all models. Models are only included in a figure if they provided results for all SSPs in that panel. There is no SSP3 in the

top row, as 1.9 Wm⁻² is infeasible in this world. Data is from an update of the IAMC Scenario Explorer developed for the SR15 (Huppmann et al. 2018; Rogelj et al. 2018a).

Implications for mitigation and other land challenges

The combination of baseline emissions development, technology options, and policy support makes it much easier to reach the climate targets in the SSP1 scenario than in the SSP3 scenario. As a result, carbon prices are much higher in SSP3 than in SSP1. In fact, the 1.9 Wm⁻² target was found to be infeasible in the SSP3 world (Cross-Chapter Box 9 Table 1). Energy system CO₂ emissions reductions are greater in the SSP3 than in the SSP1 to compensate for the higher land-based CO₂ emissions.

Accounting for mitigation and socioeconomics alone, food prices (an indicator of food insecurity) are higher in SSP3 than in the SSP1 and higher in the 1.9 Wm⁻² than in the 4.5 Wm⁻² (Cross-Chapter Box 9 Table 1). Forest cover is higher in the SSP1 than the SSP3 and higher in the 1.9 Wm⁻² than in the 4.5 Wm⁻². Water withdrawals and water scarcity are in general higher in the SSP3 than the SSP1 (Hanasaki et al. 2013a; Graham et al. 2018b) and higher in scenarios with more bioenergy (Hejazi et al. 2014c); however, these indicators have not been quantified for the specific SSP-RCP combinations discussed here.

Climate change, results in higher impacts and risks in the 4.5 Wm⁻² world than in the 1.9 Wm⁻² world for a given SSP and these risks are exacerbated in SSP3 compared to SSP1 and SSP2 due to population's higher exposure and vulnerability. For example, the risk of fire is higher in warmer worlds; in the 4.5 Wm⁻² world, the population living in fire prone regions is higher in the SSP3 (646 million) than in the SSP2 (560 million) (Knorr et al. 2016). Global exposure to multi-sector risk quadruples between the 1.5°C¹⁰ and 3°C and is a factor of six higher in the SSP3-3°C than in the SSP1-1.5°C (Byers et al. 2018). Future risks resulting from desertification, land degradation and food insecurity are lower in the SSP1 compared to SSP3 at the same level of warming. For example, the transition moderate to high risk of food insecurity occurs between 1.3 and 1.7°C for the SSP3, but not until 2.5 to 3.5°C in the SSP1 (Section 7.2).

Table 1: Quantitative indicators for the illustrative pathways. Each cell shows the mean, minimum, and maximum value across IAM models for each indicator and each pathway in 2050 and 2100. All IAMs that provided results for a particular pathway are included here. Note that these indicators exclude the implications of climate change. Data is from an update of the IAMC Scenario Explorer developed for the SR15 (Huppmann et al. 2018; Rogelj et al. 2018b).

| | | SSP1 | | SSP2 | | SSP3 | |
|---|------|--|--|--|--|--|--|
| | | 1.9 Wm ⁻² mean (min, max) | 4.5 Wm ⁻² mean (min, max) | 1.9 Wm ⁻² mean (min, max) | 4.5 Wm ⁻² mean (min, max) | 1.9 Wm ⁻² mean (min, max) | 4.5 Wm ⁻² mean (min, max) |
| Population (billion) | 2050 | 8.5 (8.5, 8.5) | 8.5 (8.5, 8.5) | 9.2 (9.2, 9.2) | 9.2 (9.2, 9.2) | N/A | 10.0 (10.0, 10.0) |
| | 2100 | 6.9 (7.0, 6.9) | 6.9 (7.0, 6.9) | 9.0 (9.0, 9.0) | 9.0 (9.1, 9.0) | N/A | 12.7 (12.8, 12.6) |
| Change in GDP per capita (% rel to 2010) | 2050 | 170.3 (380.1, | 175.3 (386.2, | 104.3 (223.4, | 110.1 (233.8, | N/A | 55.1 (116.1, 46.7) |

¹⁰ FOOTNOTE: Pathways that limit radiative forcing in 2100 to 1.9 Wm⁻² result in median warming in 2100 to 1.5°C in 2100 (Rogelj et al. 2018b). Pathways limiting radiative forcing in 2100 to 4.5 Wm⁻² result in median warming in 2100 above 2.5°C (IPCC 2014).

| | | | | | | | |
|--|------|-------------------------------|-----------------------------|--------------------------------|----------------------------|-----|----------------------------|
| | | 130.9) | 166.2) | 98.7) | 103.6) | | |
| | 2100 | 528.0 (1358.4, 408.2) | 538.6 (1371.7, 504.7) | 344.4 (827.4, 335.8) | 356.6 (882.2, 323.3) | N/A | 71.2 (159.7, 49.6) |
| Change in forest cover (Mkm ²) | 2050 | 3.4 (9.4, - 0.1) | 0.6 (4.2, - 0.7) | 3.4 (7.0, - 0.9) | -0.9 (2.9, - 2.5) | N/A | -2.4 (-1.0, - 4.0) |
| | 2100 | 7.5 (15.8, 0.4) | 3.9 (8.8, 0.2) | 6.4 (9.5, - 0.8) | -0.5 (5.9, - 3.1) | N/A | -3.1 (-0.3, - 5.5) |
| Change in cropland (Mkm ²) | 2050 | -1.2 (-0.3, - 4.6) | 0.1 (1.5, - 3.2) | -1.2 (0.3, - 2.0) | 1.2 (2.7, - 0.9) | N/A | 2.3 (3.0, 1.2) |
| | 2100 | -5.2 (-1.8, - 7.6) | -2.3 (-1.6, - 6.4) | -2.9 (0.1, - 4.0) | 0.7 (3.1, - 2.6) | N/A | 3.4 (4.5, 1.9) |
| Change in energy cropland (Mkm ²) | 2050 | 2.1 (5.0, 0.9) | 0.8 (1.3, 0.5) | 4.5 (7.0, 2.1) | 1.5 (2.1, 0.1) | N/A | 1.3 (2.0, 1.3) |
| | 2100 | 4.3 (7.2, 1.5) | 1.9 (3.7, 1.4) | 6.6 (11.0, 3.6) | 4.1 (6.3, 0.4) | N/A | 4.6 (7.1, 1.5) |
| Change in pasture (Mkm ²) | 2050 | -4.1 (-2.5, - 5.6) | -2.4 (-0.9, - 3.3) | -4.8 (-0.4, - 6.2) | -0.1 (1.6, - 2.5) | N/A | 2.1 (3.8, - 0.1) |
| | 2100 | -6.5 (-4.8, - 12.2) | -4.6 (-2.7, - 7.3) | -7.6 (-1.3, - 11.7) | -2.8 (1.9, - 5.3) | N/A | 2.0 (4.4, - 2.5) |
| Change in other natural land (Mkm ²) | 2050 | 0.5 (1.0, - 4.9) | 0.5 (1.7, - 1.0) | -2.2 (0.6, - 7.0) | -2.2 (0.7, - 2.2) | N/A | -3.4 (-2.0, - 4.4) |
| | 2100 | 0.0 (7.1, - 7.3) | 1.8 (6.0, - 1.7) | -2.3 (2.7, - 9.6) | -3.4 (1.5, - 4.7) | N/A | -6.2 (-5.4, - 6.8) |
| Carbon price (2010 USD per tCO ₂) ^a | 2050 | 510.4 (4304.0, 150.9) | 9.1 (35.2, 1.2) | 756.4 (1079.9, 279.9) | 37.5 (73.4, 13.6) | N/A | 67.2 (75.1, 60.6) |
| | 2100 | 2164.0 (35037.7, 262.7) | 64.9 (286.7, 42.9) | 4353.6 (10149.7, 2993.4) | 172.3 (597.9, 112.1) | N/A | 589.6 (727.2, 320.4) |
| Food price (Index 2010=1) | 2050 | 1.2 (1.8, 0.8) | 0.9 (1.1, 0.7) | 1.6 (2.0, 1.4) | 1.1 (1.2, 1.0) | N/A | 1.2 (1.7, 1.1) |
| | 2100 | 1.9 (7.0, 0.4) | 0.8 (1.2, 0.4) | 6.5 (13.1, 1.8) | 1.1 (2.5, 0.9) | N/A | 1.7 (3.4, 1.3) |
| Increase in Warming above pre-industrial (°C) | 2050 | 1.5 (1.7, 1.5) | 1.9 (2.1, 1.8) | 1.6 (1.7, 1.5) | 2.0 (2.0, 1.9) | N/A | 2.0 (2.1, 2.0) |
| | 2100 | 1.3 (1.3, 1.3) | 2.6 (2.7, 2.4) | 1.3 (1.3, 1.3) | 2.6 (2.7, 2.4) | N/A | 2.6 (2.6, 2.6) |
| Change in per capita demand for food, crops (% rel to 2010) ^b | 2050 | 6.0 (10.0, 4.5) | 9.1 (12.4, 4.5) | 4.6 (6.7, - 0.9) | 7.9 (8.0, 5.2) | N/A | 2.4 (5.0, 2.3) |
| | 2100 | 10.1 (19.9, 4.8) | 15.1 (23.9, 4.8) | 11.6 (19.2, - 10.8) | 11.7 (19.2, 4.1) | N/A | 2.0 (3.4, - 1.0) |
| Change in per capita demand for food, animal | 2050 | 6.9 (45.0, - 20.5) | 17.9 (45.0, - 20.1) | 7.1 (36.0, 1.9) | 10.3 (36.0, - 4.2) | N/A | 3.1 (5.9, 1.9) |

| | | | | | | | |
|---|------|-----------------------|-------------------------|------------------------|-------------------------|-----|-------------------------|
| products (% rel to 2010) ^{b,c} | 2100 | -3.0 (19.8, -27.3) | 21.4 (44.1, -26.9) | 17.0 (39.6, -24.1) | 20.8 (39.6, -5.3) | N/A | -7.4 (-0.7, -7.9) |
| AFOLU CH ₄ Emissions (% relative to 2010) | 2050 | -39.0 (-3.8, -68.9) | -2.9 (22.4, -23.9) | -11.7 (31.4, -59.4) | 7.5 (43.0, -15.5) | N/A | 15.0 (20.1, 3.1) |
| | 2100 | -60.5 (-41.7, -77.4) | -47.6 (-24.4, -54.1) | -40.3 (33.1, -58.4) | -13.0 (63.7, -45.0) | N/A | 8.0 (37.6, -9.1) |
| AFOLU N ₂ O Emissions (% relative to 2010) | 2050 | -13.1 (-4.1, -26.3) | 0.1 (34.6, -14.5) | 8.8 (38.4, -14.5) | 25.4 (37.4, 5.5) | N/A | 34.0 (50.8, 29.3) |
| | 2100 | -42.0 (4.3, -49.4) | -25.6 (-3.4, -51.2) | -1.7 (46.8, -37.8) | 19.5 (66.7, -21.4) | N/A | 53.9 (65.8, 30.8) |
| Cumulative Energy CO ₂ Emissions until 2100 (GtCO ₂) | | 428.2 (1009.9, 307.6) | 2787.6 (3213.3, 2594.0) | 380.8 (552.8, -9.4) | 2642.3 (2928.3, 2515.8) | N/A | 2294.5 (2447.4, 2084.6) |
| Cumulative AFOLU CO ₂ Emissions until 2100 (GtCO ₂) | | -127.3 (5.9, -683.0) | -54.9 (52.1, -545.2) | -126.8 (153.0, -400.7) | 40.8 (277.0, -372.9) | N/A | 188.8 (426.6, 77.9) |

^a The SSP2-19 is infeasible in two models. One of these models sets the maximum carbon price in the SSP1-19; the carbon price range is smaller for the SSP2-19 as this model is excluded there. Carbon prices are higher in the SSP2-19 than the SSP1-19 for every model that provided both simulations.

^a Food demand estimates include waste.

^b Animal product demand includes meat and dairy.

Summary

Future pathways for climate and land use include portfolios of response and policy options. Depending on the response options included, policy portfolios implemented, and other underlying socioeconomic drivers, these pathways result in different land-use consequences and their contribution to climate change mitigation. Agricultural area declines by more than 5 Mkm² in one SSP but increases by as much as 5 Mkm² in another. The amount of energy cropland ranges from nearly zero to 11 Mkm², depending on the SSP and the warming target. Forest area declines in the SSP3 but increases substantially in the SSP1. Subsequently, these pathways have different implications for risks related to desertification, land degradation, food insecurity, and terrestrial greenhouse gas fluxes, as well as ecosystem services, biodiversity, and other aspects of sustainable development.

- 1
- 2 **6.4.5 Potential Consequences of Delayed Action**
- 3 Delayed action, both in terms of overall GHG mitigation across both land and energy sectors, as well as
- 4 delayed action in implementing the specific response options outlined in this chapter, will exacerbate the
- 5 existing land challenges due to the continued impacts of climate change and socioeconomic and other
- 6 pressures; can decrease the potential of response options and increase the costs of deployment; and will
- 7 deprive communities of immediate co-benefits, among other pressures. The major consequences of
- 8 delayed action are outlined below:
- 9 *Delayed action exposes vulnerable people to continued and increasing climate impacts:* Slower or
- 10 delayed action in implementing response options exacerbates existing inequalities and impacts and will
- 11 increase the number of people vulnerable to climate change, due to population increases and increasing
- 12 climate impacts (SR15; AR 5). Future climate change will lead to exacerbation of the existing land

1 challenges, increased pressure on agricultural livelihoods, potential for rapid land degradation, and
2 millions more people exposed to food insecurity (Chapters 3, 4, 5; Schmidhuber & Tubiello 2007). Delay
3 can also bring political risks and significant social impacts, including risks to human settlements
4 (particularly in coastal areas), large-scale migration, and conflict (Barnett & Adger 2007; Hsaing et al.
5 2013). Early action reducing vulnerability and exposure can create an opportunity for a virtual circle of
6 benefits: increased resilient livelihoods, reduced degradation of land, and improved food security (Bohle
7 et al 1994).

8 *Delayed action increases requirements for adaptation:* Failure to mitigate climate change will increase
9 requirements for adaptation. For example, it is likely that by 2100 with no mitigation or adaptation, 31–69
10 million people world-wide could be exposed to flooding (Rasmussen et al., 2018; SR15; Chapter 3); such
11 outcomes could be prevented with investments in both mitigation and adaptation now. Some specific
12 response options (e.g., reduced deforestation and degradation, reduced peatland and wetland conversion)
13 prevent further detrimental effects to the land surface; delaying these options could lead to increased
14 deforestation, conversion, or degradation, serving as increased sources of GHGs and having concomitant
15 negative impacts on biodiversity and ecosystem services (section 6.2). Response options that aim at land
16 restoration and rehabilitation can serve as adaptation mechanisms for communities facing climatic
17 stresses like precipitation variability and changes in land quality, as well as providing benefits in terms of
18 mitigation.

19 *Delayed action increases response costs and reduces economic growth:* Early action on reducing
20 emissions through mitigation is estimated to result in both smaller temperature increases as well as lower
21 mitigation costs than delayed action (Sanderson et al. 2016; Luderer et al. 2013; Fujimori et al., 2016;
22 Rose et al., 2017; van Soest et al., 2017; Luderer et al., 2018). The cost of inaction to address mitigation,
23 adaptation, and sustainable land use exceeds the cost of immediate action in most countries, depending on
24 how damage functions and social cost of carbon are calculated (Dell et al. 2023; Moore & Diaz
25 2015). Costs of acting now would be one to two orders of magnitude lower than economic damages from
26 delayed action, including both damage to assets from climate impacts, as well as potentially reduced
27 economic growth, particularly in developing countries (Moore and Diaz 2015; Luderer et al. 2013; 2016).
28 Increased health costs and costs of energy (e.g. to run air-conditioners to combat increased heat waves) in
29 the US by the end of the century alone are estimated to range from 10-58% of US GDP in 2100
30 (Deschênes and Greenstone 2011).

31 Delay also increases the costs of both mitigation and adaptation actions at later dates. In models of
32 climate-economic interactions, deferral of emissions reductions now requires trade-offs leading to higher
33 costs of several orders of magnitude and risks of higher temperatures in the longer term (Luderer et al
34 2013). Further, costs of action are likely to increase over time due to the increased severity of challenges
35 in future scenarios.

36 Conversely, timely responses in implementing response options brings economic benefits. Carbon pricing
37 is one component of economic responses to encourage adoption of response options (Jakob et al. 2016),
38 but carbon pricing alone can induce higher risk in comparison to other scenarios and pathways that
39 include additional targeted sustainability measures, such as promotion of less material- and energy-
40 intensive lifestyles and healthier diets as noted in our response options (Bertram et al. 2018). While short
41 term costs of deployment of actions may increase, better attainment of a broad set of sustainability targets
42 can be achieved through these combined measures (Bertram et al. 2018).

43 There are also investments now that can lead to immediate savings in terms of avoided damages; for
44 example, for each dollar spent on DMR, countries accrue avoided disaster-related economic losses of

1 USD4 or more (Mechler 2016). While they can require upfront investment, the economic benefits of
2 actions to ensure sustainable land management, such as increased soil organic carbon, can more than
3 double the economic value of rangelands and improve crop yields (Chapter 4; section 6.2).

4 *Delayed action reduces future policy space and decreases efficacy of some response options:* The
5 potential for some response options decreases as climate change increases; for example, climate alters the
6 sink capacity for soil and vegetation carbon sequestration, reducing the potential for increased soil
7 organic carbon, afforestation, and reforestation (6.4.2). Additionally, climate change affects the
8 productivity of bioenergy crops, influencing the potential mitigation of bioenergy and BECCS (Section
9 6.4.4).

10 For response options in the supply chain, demand side management, and risk management, while the
11 consequences of delayed action are apparent in terms of continued GHG emissions from drivers, the tools
12 for response options are not made more difficult by delay and could be deployed at any time.
13 Additionally, given increasing pressures on land as a consequence of delay, some policy response options
14 may become more cost effective while others become costlier. For example, over time, land-based
15 mitigation measures like forest and ecosystem protection are likely to increase land scarcity leading to
16 higher food prices, while demand side measures, like reduced impact diets and reducing waste, are less
17 likely to raise food prices in economic models (Stevanović et al. 2017).

18 For risk management, some response options provide timely and rapidly-deployable solutions for
19 preventing further problems, such as disaster risk management and risk sharing instruments. For example,
20 early warning systems serve multiple roles in protecting lives and property and helping people adapt to
21 longer-term climate changes and can be used immediately.

22 *Delaying action can also result in problems of irreversibility of biophysical impacts and tipping points:*
23 Early action provides a potential way to avoid irreversibility - such as degradation of ecosystems that
24 cannot be restored to their original baseline - and tipping points, whereby ecological or climate systems
25 abruptly shift to a new state. Ecosystems, such as peatlands, are particularly vulnerable to irreversibility
26 because of the difficulties of rewetting to original states (Section 6.2), and dryland grazing systems are
27 vulnerable to tipping points when ground cover falls below 50%, after which productivity falls,
28 infiltration declines, and erosion increases (Chapters 3 and 4). Further, tipping points can be especially
29 challenging for human populations to adapt to, given lack of prior experience with such system shifts
30 (Kates et al. 2012; Nuttall 2012).

31 *Policy responses require lead time for implementation; delay makes this worse:* For all the response
32 options, particularly those that need to be deployed through policy implementation, there are unavoidable
33 lags in this cycle. ‘Policy lags’, by which implementation is delayed by the slowness of the policy
34 implementation cycle, are significant across many land-based, response options (Brown et al. 2019).
35 Further, the behavioural change necessary to achieve some demand-side and risk management response
36 options often takes a long time and delay only lengthens this process (Stern 1992; Steg & Vlek 2009). For
37 example, actively promoting the need for healthier and more sustainable diets through individual dietary
38 decisions is an important underpinning and enabling step for future changes, but is likely to be a slow-
39 moving process, and delay in beginning will only exacerbate this.

40 *Delay can lead to lock-in:* Delay in implementation can cause ‘lock-in’ as decisions made today can
41 constrain future development and pathways. For example, decisions made now on where to build
42 infrastructure, make investments and deploy technologies, will have longer-term (decades-long)
43 ramifications due to inertia of capital stocks (van Soest et al. 2017). In tandem, the vulnerability of the
44 poor is likely to be exacerbated by climate change creating a vicious circle of “lock in” whereby an

1 increasing share of the dwindling carbon budget may be needed to assist with improved energy use for the
2 poorest (Lamb and Rao 2015).

3 *Delay can increase the need for widespread deployment of land-based mitigation* (afforestation, BECCS)
4 (IPCC 2018; Streffler et al. 2018): Further delays in mitigation could result in an increased need for CDR
5 options later; for example, delayed mitigation requires a 10% increase in cumulative CDR over the
6 century (IPCC 2018). Similarly, strengthening near-term mitigation effort can reduce the CDR
7 requirements in 2100 by a factor of 2-8 (Streffler et al., 2018). Conversely, scenarios with limited CDR
8 require earlier emissions reductions (van Vuuren et al. 2017b) and may make more stringent mitigation
9 scenarios, like the 1.5C, infeasible (Kriegler et al. 2018a,b).

10

11 **Frequently Asked Questions**

12

13 **FAQ 6.1: What types of land-based options can help mitigate and adapt to climate change?**

14 Land-based options that help mitigate climate change are various and differ greatly in their mitigation
15 potential. The options with the moderate to large mitigation potential, and no adverse side-effects,
16 include options that decrease pressure on land (e.g. by reducing the land needed for food production) and
17 those that help to maintain or increase carbon stores both aboveground (e.g. forest measures,
18 agroforestry, fire management) and belowground (e.g. increased soil organic matter or reduced losses,
19 cropland and grazing land management, urban land management, reduced deforestation and forest
20 degradation). These options also have co-benefits for adaptation by improving health, increasing yields,
21 flood attenuation and reducing urban heat island effects. Another group of practices aim at reducing
22 greenhouse emission sources, such as livestock management or nitrogen fertilisation management. Land-
23 based options delivering climate change adaptation may be structural (e.g. irrigation and drainage
24 systems, flood and landslide control), technological (e.g. new adapted crop varieties, changing planting
25 zones and dates, using climate forecasts), or socio-economic and institutional (e.g. regulation of land use,
26 associativity between farmers). Some adaptation options (e.g. new planting zones, irrigation) may have
27 adverse-side effects for biodiversity and water. Adaptation options may be planned, such as those
28 implemented at regional, national or municipal level (top-down approaches), or autonomous, such as
29 many technological decisions taken by farmers and local inhabitants. In any case, their effectiveness
30 depends greatly on the achievement of resilience against extreme events (e.g. floods, droughts, heat
31 waves, etc.).

32 **FAQ 6.2: Which land-based mitigation measures could affect desertification, land degradation or 33 food security?**

34 Some options for mitigating climate change are based on increasing carbon stores both above and below
35 ground, so mitigation is usually related to increases in soil organic matter content and increased land
36 cover by perennial vegetation. There is a direct relationship, with very few or no adverse side-effects for
37 prevention or reversal of desertification and land degradation and the achievement of food security. This
38 is so because both desertification and land degradation are closely associated with soil organic matter
39 losses and the presence of bare ground surfaces. Food security depends on the achievement of healthy
40 crops and high and stable yields over time, which is difficult to achieve in poor soils that are low in
41 organic matter.

42 **FAQ 6.3: What is the role of bioenergy in climate change mitigation and what are its challenges?**

1 Plants absorb carbon as they grow. If plant-based material (biomass) is used for energy, the carbon it
2 absorbed from the atmosphere is released back. Traditional use of bioenergy for cooking and heating is
3 still widespread throughout the world. Modern conversion to electricity, heat, gas and liquid fuels can
4 reduce the need to burn fossil fuels and this can reduce greenhouse gas emissions, helping to mitigate
5 climate change. However, the total amount of emissions avoided depends on the type of biomass, where
6 it is grown, how it is converted to energy, and what type of energy source it displaces. Some types of
7 bioenergy require dedicated land (e.g., canola for biodiesel, perennial grasses, short rotation woody
8 crops), while others can be co-produced or use agricultural or industrial residues (e.g., residues from
9 sugar and starch crops for ethanol, manure for biogas). Depending on where, how, and the amount of
10 bioenergy crops that are grown, the use of dedicated land for bioenergy could compete with food crops
11 or other mitigation options. It could also result in land degradation, deforestation or biodiversity loss. In
12 some circumstances, however, bioenergy can be beneficial for land, for example by increasing soil
13 organic carbon. The use of co-products and residues for bioenergy limits the competition for land with
14 food but could result in land degradation if carbon and nutrient-rich material is removed that would
15 otherwise be left on the land. On the other hand, the by-products of some bioenergy conversion
16 processes can be returned to the land as a fertiliser and may have other co-benefits (e.g. reducing
17 pollution associated with manure slurry).

18

1 **Supplementary Material for Chapter 6: Interlinkages between Desertification, Land Degradation,**
 2 **Food Security and GHG fluxes: synergies, trade-offs and Integrated Response Options**

3
 4 **Supplementary Information for Section 6.4.1**

5 Section 6.4.1 includes tables of feasibility dimensions for each of the 40 response options. This section includes the supporting material for those
 6 classifications.

7 **Table SM6.1 Feasibility of land management response options in agriculture, considering cost, technological, institutional, socio-cultural and environmental and**
 8 **geophysical barriers and saturation and reversibility**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|------------------------------|------------|---------------|----------------------------------|--|---|---|--|
| Increased food productivity | | | | Limited ability to define and measure indicators of sustainable intensification (Barnes and Thomson 2014b) | better access to credit, services, inputs and markets (Schut et al. 2016) | educational (e.g., educational needs of women; Pretty and Bharucha 2014, and cultural / behavioural (Martin et al. 2015) | since increasing food productivity can be limited by climatic and environmental factors (Olesen et al. 2002) |
| Improved cropland management | | | USD74 to USD226 ha ⁻¹ | e.g., need for further development of nitrification inhibitors (Singh and Verma 2007b) | can be institutional in some regions (e.g., poor sustainability frameworks, Madlener et al. 2006) | educational (e.g., lack of knowledge; Reichardt et al. 2009b) and cultural / behavioural (e.g., promotion of cover crops needs to account for farmers' needs; Roesch-McNally et al. 2017) | e.g., land access (Bryan et al. 2009b; Bustamante et al. 2014c) |

| | | | | | | | |
|----------------------------------|--|--|--|---|---|---|--|
| Improved grazing land management | | | < USD1 kg of meat ⁻¹ (Rolfe et al., 2010) | e.g., need for further development of nitrification inhibitors (Singh and Verma 2007b) | can be institutional in some regions (e.g., need for extension services; Ndoro et al., 2014) | educational (e.g., poor knowledge of best animal husbandry practices among farmers; Ndoro et al., 2014), and cultural / behavioural (e.g. strong cultural importance of livestock and traditional practices in some communities (Herrero et al. 2016) | e.g., unless degraded, grazing lands are already closer to saturation than croplands; Smith et al. 2015 |
| Improved livestock management | | | USD120 to USD621 ha ⁻¹ (Barnhart et al., 2000) | e.g., many dietary additives are still at low technology readiness level; Beauchemin et al., 2008 | can be institutional in some regions (e.g., need for extension services; Ndoro et al., 2014), | educational (e.g., poor knowledge of best animal husbandry practices among farmers; Ndoro et al., 2014), and cultural / behavioural (e.g., strong cultural importance of livestock in some communities (Herrero et al. 2016) | e.g., climate suitability of different cattle breeds in a changing climate (Thornton et al. 2009b; Rojas-Downing et al. 2017b) |
| Agroforestry | | | < USD5 tCO ₂ e ⁻¹ (Torres et al. 2010) Note that lack of reliable financial support | There are likely to be relatively few technological barriers (Smith et al. 2007). | institutional in some regions (e.g., seed availability; (Lillesø et al. 2011) | educational (e.g., poor knowledge of how best to integrate trees into agro-ecosystems, (Meijer et al. 2015b); lack of | susceptibility to pests (Sileshi et al. 2008) |

| | | | | | | | |
|--|--|--|---|---|---|---|---|
| | | | (Hernandez-Morcillo et al. 2018) could be a barrier. | | | information, (Hernandez-Morcillo et al. 2018) and cultural / behavioural (e.g., farmers perceptions, Meijer et al. 2015b) | |
| Agricultural diversification | | | Minimal (Wimmer et al. 2016) Diversification results in cost-saving and risk reduction, thus expected cost is minimal. Note that not always economically viable (Barnes et al. 2015) | technological, biophysical, educational, and cultural barriers may emerge that limit the adoption of more diverse farming systems by farmers (Barnett and Palutikof 2015; Ahmed and Stepp 2016a); Roesch-McNally et al. 2016) | | technological, biophysical, educational, and cultural barriers may emerge that limit the adoption of more diverse farming systems by farmers (Barnett and Palutikof 2015; Ahmed and Stepp 2016a); Roesch-McNally et al. 2016) | technological, biophysical, educational, and cultural barriers may emerge that limit the adoption of more diverse farming systems by farmers (Barnett and Palutikof 2015; Ahmed and Stepp 2016a); Roesch-McNally et al. 2016) |
| Reduced grassland conversion to cropland | | | Minimal (Garibaldi et al. 2017) With increased demand for livestock products, it is expected that livestock has higher returns than crops. Note that avoiding conversion is low cost, but there | Since the response option involves not cultivating a current grassland, there are likely to be few biophysical or technological barriers | There could be institutional barriers in some regions (e.g., poor governance to prevent conversion) | educational (e.g., poor knowledge of the impacts of ploughing grasslands, and cultural / behavioural (e.g., strong cultural importance of crop production in some communities | Since the response option involves not cultivating a current grassland, there are likely to be few biophysical or technological barriers |

| | | | | | | | |
|-----------------------------|--|--|--|--|--|--|--|
| | | | may be significant opportunity costs associated with foregone production of crops. | | | | |
| Integrated water management | | | Minimal (Lubell et al. 2011) Integrated water management expected to reduce production costs and increase economic efficiency | | | | |

1

2

3

Table SM6.2 Feasibility of land management response options in forests, considering cost, technological, institutional, socio-cultural and environmental and geophysical barriers and saturation and reversibility

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|----------------------------|------------|---------------|--|---------------|---|--|--|
| Improved forest management | | | USD70 to USD160 ha ⁻¹ (Singer 2016) | | e.g., better access to credit and markets, etc. | educational (e.g., limited knowledge of the most appropriate techniques) | Forest management affects the climate also through biophysical effects and the emissions of biogenic volatile organic compounds (BVOCs), which are both influenced by species composition. |
| Reduced deforestation and | | | USD500 to USD2600 ha ⁻¹ | | e.g., land tenure, economic | educational (e.g., little information | e.g., susceptibility to climate and |

| | | | | | | | |
|--------------------------------------|--|--|--|--|--|--|--|
| degradation | | | <p>Agricultural expansion is the major driver of deforestation in developing countries. Cost of reducing of deforestation is based on opportunity cost of not growing the most common crop in developing countries (Maize) for six years to reach tree maturity, with yield of 8 t ha⁻¹ (high); 5 tons ha⁻¹ (medium) & 1.5 t ha⁻¹ & price of USD329 t⁻¹.</p> <p>Also, reduced deforestation practices have relatively moderate costs, but they requires transaction and administration costs (Overmars et al. ; Kindermann et al. 2008).</p> | | disincentives and transaction costs (Kindermann et al. 2008) | available in some regions) and cultural (different realities, e.g., small holder versus industrial production) | other unpredicted events (Ellison et al. 2017a) |
| Reforestation and forest restoration | | | USD10 to USD100 tCO ₂ e ⁻¹ (McLaren 2012b) | | | educational (e.g., low genetic diversity of planted forests) | e.g. availability of native species seedlings for planting |

| | | | | | | | |
|---------------|--|--|--|--|--|--|--|
| | | | | | | and cultural (e.g., care of forest cultures) | |
| Afforestation | | | USD10 to USD100 tCO ₂ e ⁻¹ (McLaren 2012b) | | e.g., policy makers commitment (Idris Medugu et al. 2010b) | | |

1

2

3

Table SM6.3 Feasibility of land management response options for soils, considering cost, technological, institutional, socio-cultural and environmental and geophysical barriers and saturation and reversibility

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|---------------------------------------|------------|---------------|--|---|--|---|--|
| Increased soil organic carbon content | | | USD50 to USD170 ha ⁻¹ (FAO 2014) Based on smallholder farming - which accounts for 72% farms in the world; India farmers (medium farmers) and largescale farmers in the US (FAO 2014). The cost indicated is only for manure application and ignores other costs done under business as usual (BAU). Assumes application of 10 t ha ⁻¹ of organic | e.g., difficult to measure and verify; (Smith 2006) | Can be institutional in some regions (e.g., lack of institutional capacity; Bustamante et al. 2014c) | educational (e.g., poor knowledge of best practices among farmers; (Reichardt et al. 2009b) though cultural / behavioural barriers are likely to be small compared to other barriers (Smith et al. 2007; Wollenberg et al. 2016b) | e.g., soil type; (Baveye et al. 2018b) |

| | | | | | | | |
|---------------------------|--|--|---|---|---|--|--|
| | | | manure after every three years and minimum tillage. | | | | |
| Reduced soil erosion | | | <p>USD50 to USD240 ha⁻¹ (Morokong et al. 2019)</p> <p>Based on prevention of soil erosion using terraces using rocks. Costs reported is only for avoided loss of carbon sequestration.</p> | Limited technology choices and technical support (Haregeweyn et al. 2015) | For instance, in Ethiopia farmers have shown an increased understanding of the soil erosion problem, but soil conservation programs face a host of barriers related to limited access to capital, limited benefits, land tenure insecurity (Haregeweyn et al. 2015) | Poor community participation (Haregeweyn et al. 2015) | |
| Reduced soil salinisation | | | <p>USD50 to USD250 ha⁻¹ (ICARDA 2012)</p> <p>For NENA region, salinity control recommended practice is deep ploughing, done once every 4 to 5 years to breakdown the hardpan subsoil. Deep ploughing costs USD200 ha⁻¹ for the four-year cycle or USD50 ha⁻¹ for each cropping season.</p> | e.g., lack of appropriate irrigation technology; (Machado and Serralheiro 2017b; CGIAR 2016; Bhattacharyya et al. 2015) | Lack of alternative irrigation infrastructure; (Evans and Sadler 2008; CGIAR 2016) | educational (poor knowledge of the causes and salinisation and how to address it; (Greene et al. 2016; Dagar et al. 2016b), and cultural / behavioural (persistence of traditional practices; (Greene et al. 2016; Dagar et al. 2016b) | e.g., lack of alternative water sources; (Bhattacharyya et al. 2015; Dagar et al. 2016b) |

| | | | | | | | |
|--------------------------|--|--|--|---|---|---|---|
| Reduced soil compaction | | | Negative cost (McLaren 2012b) | Both compaction process and remediation technologies are well-known (Antille et al. 2016b) but technological barriers exist (e.g., few decision support systems for implementation of precision management of traffic compaction) | | educational (knowledge gaps; Antille et al. 2016b) | Some soils are prone to compaction (Antille et al. 2016b) |
| Biochar addition to soil | | | USD100 to USD800 tCO ₂ e ⁻¹ (McLaren 2012b) A small amount of biochar potential could be available at negative cost, and some at low cost, depending on markets for the biochar as a soil amendment (Shackley et al. 2011b; Meyer et al. 2011; Dickinson et al. 2014) | e.g., feedstock and pyrolysis temperature have large impacts on biochar properties | Can be institutional in some regions (e.g., lack of quality standards; Guo et al. 2016) | educational (e.g., low awareness among end users; Guo et al. 2016) and cultural / behavioural (Guo et al. 2016) | e.g., land available for biomass production (Woolf et al. 2010) |

1
2
3
4

Table SM6.4 Feasibility of land management response options in any/other ecosystems, considering cost, technological, institutional, socio-cultural and environmental and geophysical barriers and saturation and reversibility

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|--|------------|---------------|--|--|---|--|--|
| Fire management | | | USD0.2 to USD6.5 billion per country per year (USA, Australia, Canada) | Technologies for fire management exist, but the cost of its implementation is relatively moderate, since it requires constant maintenance (North et al. 2015a) and can be excessive for some local communities. | e.g., lack of social or political acceptance (Freeman et al. 2017b) | educational (e.g., poor knowledge of best practices, liability issues, casualty risks and little tolerance for management errors; North et al. 2015a) | e.g., susceptibility to climate and other unpredicted events (Hurteau et al. 2014) or steep or remote areas to its application (North et al. 2015a) |
| Reduced landslides and natural hazards | | | | The implementation of practices for management of landslides and natural hazards is based on engineering works and more resilient cropping systems (Noble et al. 2014; Gill and Malamud 2017b), which are often limited by their high costs, as well as biophysical, technological and educational barriers. | In the tropics, the most cited barriers for implementing landslide risk reduction measures are scientific and political in nature, and the ratio of implemented versus recommended landslide risk reduction measures is low for most landslide risk reduction components (Maes et al. 2017b). | The implementation of practices for management of landslides and natural hazards is based on engineering works and more resilient cropping systems (Noble et al. 2014; Gill and Malamud 2017b), which are often limited by their high costs, as well as biophysical, technological and educational barriers. | The implementation of practices for management of landslides and natural hazards is based on engineering works and more resilient cropping systems (Noble et al. 2014; Gill and Malamud 2017b), which are often limited by their high costs, as well as biophysical, technological and educational barriers. |
| Reduced pollution including | | | USD2 to USD13 per household | e.g., lack of technology to | e.g., poor regulation and | | Since air pollution is transboundary, |

| | | | | | | | |
|--|--|--|--|---|---|--|--|
| acidification | | | (Houtven et al. 2017) | inject fertilisers below ground to prevent ammonia emissions; Shah et al., 2018 | enforcement of environmental regulations; Yamineva and Romppanen, 2017 | | sources are often far distant from the site of impact; Begum et al., 2011 |
| Management of invasive species / encroachment | | | USD500 to USD6632 per ha (Jardine et al. 2017) High cost is for California invasive alien species control; low cost from control in Massachusetts | In the case of natural enemies can be technological (Dresner et al. 2015) | Where agricultural extension and advice services are poorly developed | Education can be a barrier, where populations are unaware of the damage caused by the invasive species. Cultural / behavioural barriers are likely to be small. | Restoration programmes can take a long time (Dresner et al. 2015) |
| Restoration and reduced conversion of coastal wetlands | | | Costs for coastal wetland restoration projects vary, but they can be cost-effective at scale (Erwin 2009) | | Can be institutional in some regions (e.g., poor governance of wetland use in some regions; (Lotze et al. 2006) | educational (e.g., lack of knowledge of impact of wetland conversion), though technological and cultural / behavioural barriers are likely to be small compared to other barriers. | e.g., loss of large predators, herbivores, spawning and nursery habitat; (Lotze et al. 2006) |
| Restoration and reduced conversion of peatlands | | | USD4 to USD20 tCO ₂ e ⁻¹ (McLaren 2012b) | | An be institutional in some regions (e.g., lack of inputs; Bonn et al. 2014) | educational (e.g., lack of skilled labour; Bonn et al. 2014), though technological and cultural / behavioural barriers are likely to be small | e.g., site inaccessibility; Bonn et al. 2014) |

| | | | | | | | |
|---------------------------|--|--|--|--|--|-----------------------------|--|
| | | | | | | compared to other barriers. | |
| Biodiversity conservation | | | USD10 to USD50 tCO ₂ e ⁻¹ (Minx et al. 2018) | | | | |

1
2
3
4

Table SM6.5 Feasibility of land management response options specifically for CDR, considering cost, technological, institutional, socio-cultural and environmental and geophysical barriers and saturation and reversibility

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|---------------------------------|------------|--|--|---|---|---|--|
| Enhanced weathering of minerals | | | USD10 to USD40 tCO ₂ e ⁻¹ (McLaren 2012b) The main cost (and large energy input) is in the mining and comminution of the minerals (Renforth et al. 2012) with higher total costs compared to other low cost land management options (Smith et al. 2016a). | High energy costs of comminution; Smith et al. 2016a | In some regions (e.g., lack of infrastructure for this new technology; Taylor et al. 2016c) | Educational (e.g., lack of knowledge of how to use these new materials in agriculture). Cultural barriers could occur in some regions, for example, due to minerals lying under undisturbed natural areas where mining might generate public acceptance issues (e.g., Renforth et al. 2012) | e.g., limited and inaccessible mineral formations (Renforth et al. 2012) |
| Bioenergy and BECCS | | BECCS "is one of the NET options that is less vulnerable to reversal" (Fuss et al. 2018) | USD50 to USD250 tCO ₂ e ⁻¹ (McLaren 2012b) | While there are a few small BECCS demonstration facilities, BECCS has not been implemented at | Institutional barriers include governance issues (Gough 2016) | Cultural barriers include social acceptance (Sanchez and Kammen 2016b) with CCS facing | Competition for land and water |

| | | | | | | | |
|--|--|--|--|----------------------|--|---|--|
| | | | | scale (Kemper 2015b) | | concerns of safety and environmental issues and bioenergy facing additional scrutiny because of competition for land and water. | |
|--|--|--|--|----------------------|--|---|--|

1

2 **Table SM6.6 Feasibility of demand management response options, considering cost, technological, institutional, socio-cultural and environmental and geophysical**
 3 **barriers and saturation and reversibility**

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|-----------------------------|------------|---------------|------|--|--|--|--|
| Dietary change | | | | Inadequate storage options for e.g. fresh fruit and vegetables | Barriers might also be institutional in some regions (e.g., poorly developed dietary health advice, Wardle et al. 2000b) | cultural / behavioural (e.g., diets are deeply culturally embedded and behaviour change is extremely difficult to effect, even when health benefits are well known; Macdiarmid et al., 2016); educational (e.g., poor knowledge of what constitutes a healthy diet; Wardle et al. 2000b) | poor accessibility of healthy foods such and fruit and vegetables (e.g., Hearn et al. 1998b; Lock et al. 2005) |
| Reduced post-harvest losses | | | | Lack of low-cost storage and preservation technologies | Barriers are largely institutional, since solutions may require | There are few biophysical, educational or cultural barriers, since preventing | There are few biophysical, educational or cultural barriers, since preventing |

| | | | | | | | |
|---|--|--|--|---|---|--|---|
| | | | | | dismantling and redesigning current food value chains | food loss is a priority in many developing countries. | food loss is a priority in many developing countries. |
| Reduced food waste (consumer or retailer) | | | | Barriers in developing countries include reliability of transportation networks, market reliability, education, technology, capacity, and infrastructure (Kummu et al. 2012). | Specific barriers to reducing consumption waste in industrialised countries include inconvenience, lack of financial incentives, lack of public awareness, low cost of food, quality standards and regulations, consumer's ability to buy food products at any time, generalised oversupply in the distribution, and low prioritisation, among others (Kummu et al.); (Graham-Rowe et al. 2014); Diaz-Ruiz et al., 2018). Barriers in developing countries include reliability of transportation networks, market reliability, education, technology, capacity, and infrastructure (Kummu et al.) | Specific barriers to reducing consumption waste in industrialised countries include inconvenience, lack of financial incentives, lack of public awareness, and low prioritisation (Kummu et al.); (Graham-Rowe et al. 2014). Barriers in developing countries include reliability of transportation networks, market reliability, education, technology, capacity, and infrastructure (Kummu et al.) | |

| | | | | | | | |
|-----------------------|--|--|----------------------------|---|--|--|--|
| | | | | | infrastructure (Kummu et al.) | | |
| Material substitution | | | Negligible (McLaren 2012b) | Improved treatments to prevent against fire and moisture needed (Ramage et al. 2017b) | Construction companies hesitant to take risks associated with wooden buildings and insurance companies rate wooden buildings as higher risk (Gustavson et al., 2006) | People perceive adverse effects of wood products on forests and increased risk of fire (Gustavson et al. 2006) | |

1

2

3

4

Table SM6.7 Feasibility of supply management response options, considering cost, technological, institutional, socio-cultural and environmental and geophysical barriers and saturation and reversibility

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|----------------------|---|---|------|---------------|---|---|---|
| Sustainable sourcing | Reversibility could be an issue and while there are low cost options, the implementations can be expensive. | Reversibility could be an issue and while there are low cost options, the implementations can be expensive. | | | There are institutional barriers in some contexts (e.g., in low income African, Asian and Latin American countries where challenges associated with food insecurity and climate change vulnerability are more acute) (Ingram et al. | No obvious biophysical or cultural barriers | No obvious biophysical or cultural barriers |

| | | | | | | | |
|--|--|--|---|--|---|--|--|
| | | | | | 2016a) | | |
| Management of supply chains | | | | | political will within trade regimes, economic laissez-faire policies that discourage interventions in markets, and the difficulties of coordination across economic sectors (Poulton et al. 2006; Cohen et al. 2009; Gilbert 2012b) | | |
| Enhanced urban food systems | | | | There are likely to be few biophysical, technological or cultural / behavioural barriers to implementing improved urban food systems, though institutional and education barriers could play a role. | There are likely to be few biophysical, technological or cultural / behavioural barriers to implementing improved urban food systems, though institutional and education barriers could play a role. | There are likely to be few biophysical, technological or cultural / behavioural barriers to implementing improved urban food systems, though institutional and education barriers could play a role. | There are likely to be few biophysical, technological or cultural / behavioural barriers to implementing improved urban food systems, though institutional and education barriers could play a role. |
| Improved food processing and retailing | | | The implementation of strategies to improve the efficiency and sustainability of retail and agri-food industries can be expensive | Adoption of specific sustainability instruments and eco-innovation practices | Successful implementation is dependent on organisational capacity, the agility and flexibility of business strategies, the | No obvious cultural/behavioural barriers, but educational barriers exist | No obvious biophysical and cultural/behavioural barriers |

| | | | | | | | |
|-------------------------------------|--|--|--|--|--|--|--|
| | | | | | strengthening of public-private policies and effectiveness of supply-chain governance. | | |
| Improved energy use in food systems | | | | e.g., low levels of farm mechanisation | e.g., energy efficiency in agriculture depends strongly on the technology level (Vlontzos et al. 2014) | educational (e.g., poor knowledge of alternative energy sources), and behavioural / cultural (e.g., high levels of repetitive labour, making farming unattractive to the youth, and disproportionately affecting women; (Baudron et al. 2015b) | |

1
2
3
4

Table SM6.8 Feasibility of risk management response options, considering cost, technological, institutional, socio-cultural and environmental and geophysical barriers and saturation and reversibility

| Response option | Saturation | Reversibility | Cost | Technological | Institutional | Socio-cultural | Environmental and geophysical |
|----------------------------|------------|---------------|--|---------------|--|----------------|-------------------------------|
| Management of urban sprawl | | | USD0.5 to USD3 trillion yr ⁻¹ globally (New Climate Economy 2018) Global cost of prevention of urban sprawl done | | Barriers to policies against urban sprawl include institutional barriers to integrated land use planning and the costs to national | | |

| | | | | | | |
|----------------------------|--|--|---|--|--|---|
| | | | by: densification; provision of sustainable and affordable housing; and investment in shared, electric, and low-carbon transport. | | governments of restricting or buying back development rights (Tan et al. 2009) | |
| Livelihood diversification | | | Barriers to diversification include the fact that poorer households and female headed households may lack assets to invest in new income streams or have a lack of education about new income sources (Berman et al. 2012b; Ahmed and Stepp 2016a; Ngigi et al. 2017) | | | Barriers to diversification include the fact that poorer households and female headed households may lack assets to invest in new income streams or have a lack of education about new income sources (Berman et al. 2012b; Ahmed and Stepp 2016a; Ngigi et al. 2017) |
| Use of local seeds | | | | | | Barriers to seed sovereignty include concerns about equitability in access to seed networks and the difficulty of sustaining such projects when development donors leave (Reisman 2017b), |

| | | | | | | | |
|--------------------------|--|--|--|--|---|---|--|
| | | | | | | and disputes over the intellectual property rights associated with seeds (Timmermann and Robaey 2016) | |
| Disaster risk management | | | Barriers to EWS include cost; an early warning system for the 80 most climate vulnerable countries in the world is estimated to cost USD 2 billion over five years to develop (Hallegatte 2012). | | Institutional and governance barriers such as coordination and synchronisation among levels also effect some EWS (Birkmann et al. 2015b). | | |
| Risk sharing instruments | | | USD10 to USD90 ha ⁻¹ (Schmitkey 2017) Insurance cost depends on value of crop. We use maize as an example in US (high) and Sub-Saharan Africa (low). | | | | |

1

1 **Supplementary Information for Section 6.4.3**

2 Section 6.4.3 includes tables regarding interactions for each of the 40 response options with Nature’s Contributions to People (NCP) and Sustainable
 3 Development Goals (SDG). This section includes the supporting material for those classifications.

4 **Table SM6.9 Impacts on Nature’s Contributions to People of integrated response options based on land management**

| <u>Integrated response options based on land management</u> | Habitat creation and maintenance | Pollination and dispersal of seeds and other propagules | Regulation of air quality | Regulation of climate | Regulation of ocean acidification | Regulation of freshwater quantity, flow and timing | Regulation of freshwater and coastal water quality | Formation, protection and decontamination of soils and sediments | Regulation of hazards and extreme events | Regulation of organisms detrimental to humans | Energy | Food and feed | Materials and assistance | Medicinal, biochemical and genetic resources | Learning and inspiration | Physical and psychological experiences | Supporting identities | Maintenance of options | |
|---|--|--|---------------------------|---|---|---|--|--|--|---|--------|---|---|--|--------------------------|--|-----------------------|---|-----|
| Increased food productivity | Higher productivity spares land (e.g. Balmford et al. 2018) especially if intensification is done sustainably. | Likely may reduce native pollinators if reliant on increased chemical inputs (Potts et al. 2010) but not if through sustainable intensification. | N/A | N/A | Increased food productivity might be achieved through increased pesticide or fertiliser use, which causes runoff and dead zones in oceans (Beusen et al. 2016). | Food productivity increases could impact water quality if increases in chemicals used, but evidence is mixed on sustainable intensification (Rockström et al. 2009; Mueller et al. 2012). | Food productivity increases could impact water flow due to demand for irrigation (Rockström et al. 2009; Mueller et al. 2012). | Intensification through additional input of nitrogen fertiliser can result in negative impacts on climate, soil, water and air pollution (Tilman et al. 2002). | N/A | Increasing food production through agrochemicals may increase pest resistance over time (Tilman et al. 2002). | N/A | Sustainable intensification has potential to close yield gaps (Tilman et al. 2011). | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
| Improved cropland management | Improved cropland management can contribute to diverse agroecosystems (Tscharntke et al. 2005) and promotes soil biodiversity (Oehl et al. 2017) | Better crop management can contribute to maintaining native pollinators (Gardiner et al. 2009). | N/A | See main text for mitigation potentials | Mitigation potential (see main text) will reduce ocean acidification. | Cropland conversion has major impacts on water quantity (Scanlon et al. 2007). Cropland management practices such as conservation tillage improve downstream water quality due to runoff (Scanlon et al. 2007). | Cropland conversion leads to poorer water quality due to runoff (Scanlon et al. 2007). | Improved cropland management has positive impacts on soils (see main text) (Kern et al. 2003). | N/A | Some forms of improved cropland management can decrease pathogens and pests (Tscharntke et al. 2016). | N/A | Conservation agriculture contributes to food productivity and reduces food insecurity (Rosegrant and Cline 2003 ; Dar & Gowda 2011; Godfrey & Garnett 2014) | N/A | N/A | N/A | N/A | N/A | Many cropping systems have cultural components (Tenberg et al. 2012). | N/A |
| Agriculture Improved grazing land management | Can contribute to improved habitat (Pons et al. 2003; Plantureux et al. 2005). | N/A | N/A | See main text for mitigation potentials | Mitigation potential (see main text) will reduce ocean acidification. | Likely will improve water quality (Hibbert 1983). | Likely will improve water flow (Hibbert 1983). | Improved grassland management increases soil carbon and quality (Conant et al. 2001). | N/A | N/A | N/A | Improved grassland management could contribute to food security (O'Mara 2012) | Grassland management can provide other materials (e.g. biofuel materials) | N/A | N/A | N/A | N/A | Many pastoralists have close cultural connections to livestock (Ainslie 2013) | N/A |

| | | | | | | | | | | | | | | | | | |
|--|---|--|---|---|--|--|---|---|---|---|---|--|------------|------------|------------|------------|--|
| <p>Avoidance of conversion of grassland to cropland</p> | <p>Can preserve natural habitat (Peeters, 2009)</p> | <p>N/A</p> | <p>N/A</p> | <p>See main text for mitigation potentials</p> | <p>Mitigation potential (see main text) will reduce ocean acidification.</p> | <p>Will likely improve water quality (inferred from improved soil quality in Saziozzi et al., 2001)</p> | <p>Will likely improve water flow (inferred from improved soil quality in Saziozzi et al., 2001)</p> | <p>Will improve soil quality (Saziozzi et al., 2001)</p> | <p>N/A</p> | <p>Diverse agroecosystems tend to have less detrimental impacts from pests (Gardiner et al 2009; Altieri & Letourneau 1982)</p> | <p>N/A</p> | <p>Reducing cropland conversion can reduce food production (West et al. 2010).</p> | <p>N/A</p> | <p>N/A</p> | <p>N/A</p> | <p>N/A</p> | <p>Retaining natural ecosystems can preserve genetic diversity (Ekins et al., 2003).</p> |
| <p>Integrated water management</p> | <p>Ecosystem health and services can be enhanced by improving water management (Boelee E and E 2011). Securing ecosystem (Lloyd et al. 2013), integrated ecosystem-based management into water resources planning and management, linking ecosystem services and water security (Nicole Bernex 2016), improving correlation between amount of water resources and supply ecosystem services, combining water resources management and supply of ecosystem services (Liu et al. 2016).</p> | <p>Some integrated water management strategies generate synergies between multiple ecosystem services, such as pollination, farm profitability (Hipólito et al, 2018).</p> | <p>IWM practices exert strong influence on ecosystem structure and function, with potentially large implications for regulating air quality (Xia et al., 2017; Hardiman et al, 2019).</p> | <p>IWM supports favourable forests conditions thereby influencing the storage and flow of water in watersheds (Eisenbies et al. 2007) which are important for regulating microclimates (Pierzynski et al., 2017).</p> | <p>N/A</p> | <p>Improving regulations for water sharing, trading and pricing (ADB 2016), water smart appliance, water smart landscapes (Dawadi and Ahmad 2013), common and unconventional water sources in use (Rengasamy 2006) will increase water quantity.</p> | <p>Improving regulation to prevent aquifer and surface water depletion, controlling over water extraction, improvement of water management and management of landslides and natural hazards. Watering shifting sand dunes (sprinkler), water resources conservation (Nejad 2013; Pereira 2002a), enhancing rainwater management, reducing recharge and increasing water use in discharge areas (DERM 2011).</p> | <p>IWM provide co-benefits such as healthier soils, more resilient and productive ecosystems (Grey and Sadoff 2007; Liu et al. 2017; Scott et al. 2011)</p> | <p>Change in water availability through improving co-managing floods and groundwater depletion at the river basin such as Managed Aquifer Recharge (MAR), Underground Taming of Floods for Irrigation (UTFI), restore over-allocated or brackish groundwater dependent ecosystems protection, reducing evaporation losses are significantly contributed to response climate change and reduced impacts of extreme weather event in desertification areas (Dillon and Arshad 2016b).</p> | <p>IWM can support the production of biomass and firewood (Mbow et al., 2014).</p> | <p>Increasing demand for food, fiber and feed will put great strains on land, water, energy and other resources (WBCSD, 2014). Water conservation and balance in the use of natural resources enforcement (based water resources, water conservation measures, water allocations) (Ward et al. 2008) are good options to response climate change and nature's prevention.</p> | <p>IWM supports favourable forests conditions thereby providing wood and fodder and other materials (Locate Ili et al. 2015a). However, conservation restrictions on the storage and flow of water in watersheds (Eisenbies et al. 2007) can restrict the access to resources (e.g. firewood).</p> | | | | | |

| | | | | | | | | | | | | | | | | | | | |
|----------------|---|--|--|--|--|--|--|--|--|--|--|---|---|---|---|--|--|--|--|
| <p>Forests</p> | <p>Forest management and forest restoration</p> | <p>Forest landscape restoration specifically aims to regain ecological integrity and enhance human well-being in deforested or degraded forest landscape (Maginnis and Jackson 2007; Stanturf et al. 2014). For example, facilitating tree species mixture means storing at least as much carbon as monocultures while enhancing biodiversity (Hulvey et al. 2013). Selective logging techniques are “middle way” between deforestation and total protection, allowing to retain substantial levels of biodiversity, carbon, and timber stocks (Putz et al. 2012),</p> | <p>Likely contributes to native pollinators (Kremen et al. 2007)</p> | <p>Trees remove air pollution by the interception of particulate matter on plant surfaces and the absorption of gaseous pollutants through the leaf stomata. Computer simulations with local environmental data reveal that trees and forests in the conterminous United States removed 17.4 million tonnes (t) of air pollution in 2010 (range: 9.0–23.2 million t), with human health effects valued at 6.8 billion U.S. dollars (range: USD1.5–13.0 billion) (Novak et al., 2014)</p> | <p>See main text for mitigation potentials</p> | <p>Mitigation potential (see main text) will reduce ocean acidification.</p> | <p>Forest cover can stabilise intense runoff during storms and flood events (Locatelli et al. 2015a). Mangroves can protect coastal zones from extreme events (hurricanes) or sea level rise. However, forests also can have adverse side-effects for reduction of water yield and water availability for human consumption (Bryan and Crossman 2013).</p> | <p>Forests tend to maintain water quality by reducing runoff and trapping sediments and nutrients (Idris Medugu et al. 2010a; Salvati et al. 2014). Precipitation filtered through forested catchments delivers purified ground and surface water (co-benefits) (Calder 2005; Ellison et al. 2017; Neary et al. 2009).</p> | <p>Forests counteract wind-driven degradation of soils, and contribute to soil erosion protection and soil fertility enhancement for agricultural resilience (Locatelli et al. 2015a).</p> | <p>Forest cover can stabilise land against catastrophic movements associated with wave action and intense runoff during storms and flood events (Locatelli et al. 2015a). Reducing harvesting rates and prolonging rotation periods may induce an increased vulnerability of stands to external disturbances and catastrophic events (Yousefpour et al. 2018). Forest management strategies may decrease stand-level structural complexity and may make forest ecosystems more susceptible to natural disasters like wind throws, fires, and diseases (Seidl et al. 2014).</p> | <p>Forests can contribute to weed and pest control and landscape diversity generally improves opportunities for biological pest control (Gardiner et al. 2009)</p> | <p>SFM may increase availability of biomass for energy (Kraxner et al 2003; Sikkema et al 2014)</p> | <p>The proximity of forest to cropland constitutes a threat to livelihoods in terms of crop raiding by wild animals and in constraints in availability of land for farming (Few et al. 2017). The competition for land between afforestation/reforestation and agricultural production is a potentially large adverse side-effect (Boysen et al. 2017a,b; Kreidenweis et al. 2016; Smith et al. 2013). An increase in global forest area can lead to increases in food prices through increasing land competition (Calvin et al. 2014; Kreidenweis et al. 2016; Reilly et al. 2012; Smith et al. 2013; Wise et al. 2009).</p> | <p>Forests provide wood and fodder and other materials (Locatelli et al. 2015a). However, conservation restrictions to preserve ecosystem integrity can restrict the access to resources (e.g. firewood).</p> | <p>Can provide medicinal and other resources.</p> | <p>Natural ecosystems often inspire learning (Turtle et al., 2015)</p> | <p>Forest landscape restoration specifically aims to enhance human well-being (Maginnis and Jackson 2007; Stanturf et al. 2014). Afforestation/reforestation and avoided deforestation benefit biodiversity and species richness, and generally improve the cultural and recreational value of ecosystems (co-benefits) (Knoke et al. 2014).</p> | <p>Many forest landscapes have cultural ecosystem services components (Plieninger et al. 2015)</p> | <p>Retaining natural ecosystems can preserve genetic diversity (Ekins et al., 2003).</p> |
|----------------|---|--|--|--|--|--|--|--|--|--|--|---|---|---|---|--|--|--|--|

| | | | | | | | | | | | | | | | | | | | |
|-------|---------------------------------------|---|-----|-----|---|---|---|--|--|--|--|---|---|---|-------------------------------------|-----|---|---|-----|
| | | 2016; Ellison et al. 2017). | | | | | | | | | | | Smith et al. 2013). An increase in global forest area can lead to increases in food prices through increasing land competition (Calvin et al. 2014; Kreidenweis et al. 2016; Reilly et al. 2012; Smith et al. 2013; Wise et al. 2009). | d). | | | | | |
| | Afforestation | Forest landscape restoration specifically aims to regain ecological integrity and enhance human well-being in deforested or degraded forest landscape (Maginnis and Jackson 2007; Stanturf et al. 2014). In the case of afforestation, simply changing the use of land to planted forests is not sufficient to increase abundance of indigenous species, as they depend on type of vegetation, scale of the land transition, and time required for a population to establish (Barry et al. 2014). | N/a | N/A | See main text for mitigation potentials | Mitigation potential (see main text) will reduce ocean acidification. | Depends on where reforestation and with what species (Scott et al. 2005). Trees enhance soil infiltration and, under suitable conditions, improve groundwater recharge (Calder 2005; Ellison et al. 2017; Neary et al. 2009). | Afforestation using some exotic species can upset the balance of evapotranspiration regimes, with negative impacts on water availability particularly in arid regions (Ellison et al. 2017; Locatelli et al. 2015a; Trabucco et al. 2008). Afforestation in arid and semiarid regions using species that have evapotranspiration rates exceeding the regional precipitation may aggravate the groundwater decline (Locatelli et al. 2015a; Lu et al. 2016). Changes in runoff affect water supply but can also contribute to changes in flood risks, and irrigation of forest plantations can increase water consumption (Sterling et al. 2013). | Afforestation and reforestation options are frequently used to counteract land degradation problems (Yirdaw et al. 2017), whereas when they are established on degraded lands they are instrumental to preserve natural forests (co-benefit) (Buongiorno and Zhu 2014). Afforestation runs the risk of decreasing soil nutrients, especially in intensively managed plantations; in one study, afforestation sites had lower soil P and N content (Berthrong et al. 2009). | Some afforestation may make forest ecosystems more susceptible to natural disasters like wind throws, fires, and diseases (Seidl et al. 2014). | N/A | Afforestation may increase availability of biomass for energy use (Obersteiner et al. 2006) | Future needs for food production are a constraint for large-scale afforestation plans (Locatelli et al. 2015a). Global food crop demand is expected by 50%-97% between 2005 and 2050 (Valin et al. 2014). Future carbon prices will facilitate deployment of afforestation projects at expenses of food availability (adverse side-effect), but more liberalised trade in agricultural commodities could buffer food price increases following afforestation in tropical regions (Kreidenweis et al. 2016). | Could increase availability of biomass (Griscom et al., 2017) | N/A | N/A | Green spaces support psychological wellbeing (Coldwell & Evans, 2018) | Afforestation/ reforestation can increase areas available for recreation and tourism opportunities (Knoke et al. 2014). | N/A |
| Soils | Increased soil organic carbon content | Improving soil carbon can increase overall resilience of | N/A | N/A | See main text for mitigation potentials | Rivers transport dissolved organic | Soil organic matter is known to increase water | Soil organic matter is known to increase water filtration and | Increasing SOM contributes to healthy soils (Lehmann & | N/A | Increased SOM decreases pathogens in soil (Lehmann | N/A | Lal 2006 notes that "Food-grain production in | In terms of raw material | In terms of raw materials, numerous | N/A | N/A | N/A | N/A |

| | | | | | | | | | | | | | | | | | | | |
|--|----------------------------------|--|-----|---|-----|---|---|---|---|--|----------------|--|--|--|-----|-----|-----|-----|-----|
| | | landscapes (Tscharnke et al. 2005) | | | | matter to oceans (Hedges et al 1997), but unclear if improved SOM will decrease this and by how much. | filtration and can regulate downstream flows (Keesstra et al., 2016) | protects water quality (Lehmann & Kleber 2015) | Kleber 2015) | | & Kleber 2015) | | developing countries can be increased by 24–39 (32+/-11) million Mgy-1 through improving soil quality by increasing the SOC pool and reversing degradation processes". | s, numerous products (e.g. pharmaceuticals, clay for bricks and ceramics, silicon for bricks and electronics, and other minerals; SSSA, 2015) are provided by soils. | | | | | |
| | Reduced soil erosion | Managing soil erosion decreases need for expanded cropland into habitats (Pimental et al 1995) | N/A | Particulate matter pollution, a main consequence of wind erosion, imposes severe adverse impacts on materials, structures and climate which directly affect the sustainability of urban cities (Al-Thani et al. 2018) | N/A | N/A | Managing soil erosion improves water quality (Pimental et al 1995) | Managing soil erosion improves water flow (Pimental et al 1995) | Will improve soil quality (Keesstra et al., 2016) | | | Reducing soil erosion reduces vulnerability to hazards like wind storms in dryland areas and landslides in mountainous areas (EL-Swify 1997) | Managing erosion can lead to increased food production on croplands; however, other forms of management (revegetation, zero tillage) might reduce land available for food. | N/A | N/A | N/A | N/A | N/A | N/A |
| | Reduced soil salinisation | Salinisation decreases soil microbial diversity (Nic et al. 2009) | N/A | N/A | N/A | N/A | Management of soil salinity improves water quality (Kotb et al. 2000; Zalidis et al 2002; Soane & Ouwerkerk 1995) | | Will improve soil quality (Keesstra et al., 2016) | | | N/A | Reversing degradation contributes to food productivity and reduces food insecurity (Pimental et al. 1995; Shiferaw & Holden 1999). | N/A | N/A | N/A | N/A | N/A | N/A |

| | | | | | | | | | | | | | | | | | | | |
|-------------------------|--|--|--|---|---|---|---|--|---|--|---|--|--|-----|-----|-----|--|-----|---|
| | Reduced soil compaction | Preventing compaction can reduce need to expand croplands (Lal, 2001). | N/A | N/A | N/A | N/A | Compaction can increase water runoff (Soane & Ouwerkerk 1995). Management of soil compaction improves water quality and quantity (Soane & van Ouwerkerk 1995; Zalidis et al 2002) | Management of soil compaction improves water quality and quantity (Soane & van Ouwerkerk 1995; Zalidis et al 2002) | Will improve soil quality (Keesstra et al., 2016) | Compaction in soils increases rates of runoff and can contribute to floods (Hümann et al 2011) | N/A | N/A | Compactions reduces agricultural productivity and thus contributes to food insecurity (Navaz et al 2013) | N/A | N/A | N/A | N/A | N/A | N/A |
| | Biochar addition to soil | N/A | N/A | N/A | See main text for mitigation potentials | Mitigation potential (see main text) will reduce ocean acidification. | Biochar improves soil water filtration and retention (Spokas et al 2011; Beck et al. 2011) | Biochar improves soil water filtration and retention (Spokas et al 2011; Beck et al. 2011) | Can improve soil quality (Sohi, 2012) | N/A | N/A | N/A | Contributes to increased food production (Smith 2016; Jeffery et al., 2017) | N/A | N/A | N/A | N/A | N/A | N/A |
| | Fire management | Proactive fire management can improve natural habitat (Burrows 2008). | Reducing fire risk can improve habitat for pollinators (Brown et al. 2017) | Fire management improves air quality particularly in the periurban interface (Bowman et al. 2005) | See main text for mitigation potentials | Mitigation potential (see main text) will reduce ocean acidification. | Fires affect water quality and flow due to erosion exposure (Townsend & Douglas 2000). | Fires affect water quality and flow due to erosion exposure (Townsend & Douglas 2000). | Fire cause damage to soils, therefore fire management can improve them (Certini 2005) | Will reduce risk of wildfires as a hazard (McCaffrey 2002) | Landscape diversity generally improves opportunities for biological pest control (Gardiner et al. 2009) | Will increase availability of biomass, as fuel removal is a key management strategy (Becker et al. 2009) | N/A | N/A | N/A | N/A | Reduced wildlife risk will increase recreation opportunities in landscapes (Venn & Calkin 2011). | N/A | Retaining natural ecosystems can preserve genetic diversity (Ekins et al., 2003). |
| | Reduced landslides and natural hazards | Can preserve natural habitat (Dolidon et al. 2009) | N/A | N/A | N/A | N/A | Likely will improve water quality (Dolidon et al. 2009) | Likely will improve water flow (Dolidon et al. 2009) | Will improve soil quality (Keesstra et al., 2016) | Will reduce risk of disasters (Dolidon et al. 2009; Kausky 2010) | N/A | N/A | Landslides are one of the natural disasters that have impacts on food security (de Haen & Henrich 2007) | N/A | N/A | N/A | N/A | N/A | N/A |
| Other ecosystems | Reduced pollution including acidification | Air pollution like acid rain has major impacts on habitats like lakes (Schindler et al 1989) | Pollution interferes with scents, which impact pollinators ability to detect resources (McFredrick et al 2008) | Will improve air quality with public health benefits (Nemet et al. 2010) | See main text for mitigation potentials | N/A | N/A | Pollution increases acidity of surface water, with likely ecological effects (Larsen et al 1999) | Soil acidification due to air pollution in a serious problem in many countries (Zhou et al. 2013) | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |

| | | | | | | | | | | | | | | | | | | |
|--|---|--|------------|--|--|--|--|---|--|--|--|---|--|--|--|--|--|---|
| <p>Management of invasive species / encroachment</p> | <p>Improved management of IAS can lead to improved habitat and ecosystems (Richardson & van Wilgen 2004).</p> | <p>Invasive species can disrupt native plant-pollinator relations (Ghazoul 2006)</p> | <p>N/A</p> | <p>N/A</p> | <p>N/A</p> | <p>Many invasives can reduce water flow (Richardson & Van Wilgen 2004).</p> | <p>Invasive species can reduce water quality (Burnett et al. 2007; Chamier et al. 2012)</p> | <p>Likely to improve soil as invasive species generally have negative effects (Ehrenfeld & Scott 2001).</p> | <p>N/A</p> | <p>Many IAS are harmful pests (Charles & Dukes 2008).</p> | <p>N/A</p> | <p>IAS can compete with crops and reduce crop yields by billions of dollars annually (Pejchar & Mooney 2009)</p> | <p>Many invasives are important suppliers of materials (Pejchar & Mooney 2009).</p> | <p>N/A</p> | <p>N/A</p> | <p>N/A</p> | <p>N/A</p> | <p>Reducing invasives can increase biological diversity of native organisms (Simberloff 2005)</p> |
| <p>Restoration and avoided conversion of coastal wetlands</p> | <p>Will preserve natural habitat (Griscom et al., 2017)</p> | <p>Will promote natural pollinators (Seddon et al., 2016)</p> | <p>N/A</p> | <p>See main text for mitigation potentials</p> | <p>Mitigation potential (see main text) will reduce ocean acidification.</p> | <p>The creation or restoration of wetlands, tidal marshes, or mangroves provide water retention and protect coastal cities from storm surge flooding and shoreline erosion during storms. Wetlands store freshwater and enhance water quality (Bobbink et al 2006)</p> | <p>Wetlands store freshwater and enhance water quality (Bobbink et al 2006)</p> | <p>Will improve soil quality (Griscom et al., 2017)</p> | <p>The creation or restoration of wetlands, tidal marshes, or mangroves provide water retention and protect coastal cities from storm surge flooding and shoreline erosion during storms (Haddad et al., 2015; Gittman et al. 2014; Kaplan et al. 2009).</p> | <p>Landscape diversity generally improves opportunities for biological pest control (Gardiner et al. 2009)</p> | <p>N/A</p> | <p>Mixed evidence: can affect agriculture/fisheries production when competition for land occurs, or could increase food production when ecosystems are restored (Crooks et al 2011)</p> | <p>Could increase availability of biomass (Griscom et al., 2017)</p> | <p>Wetlands can be sources of medicines (UNEP, 2016)</p> | <p>Natural ecosystems often inspire learning (Turtle et al., 2015)</p> | <p>Natural environments support psychological wellbeing (Coldwell & Evans, 2018)</p> | <p>Natural environments support psychological wellbeing (Coldwell & Evans, 2018)</p> | <p>Retaining natural ecosystems can preserve genetic diversity (Ekins et al., 2003).</p> |
| <p>Restoration and avoided conversion of peatlands</p> | <p>Will preserve natural habitat (Griscom et al., 2017)</p> | <p>Could promote natural pollinators (Seddon et al., 2016)</p> | <p>N/A</p> | <p>See main text for mitigation potentials</p> | <p>Mitigation potential (see main text) will reduce ocean acidification.</p> | <p>Peatland restoration will improve water quality as they play important roles in water retention and drainage (Johnston 1991).</p> | <p>Peatland restoration will improve water quality as they play important roles in water retention and drainage (Johnston 1991).</p> | <p>Will improve soil quality (Griscom et al., 2017)</p> | <p>N/A</p> | <p>Landscape diversity generally improves opportunities for biological pest control (Gardiner et al. 2009)</p> | <p>Will reduce supply of any biomass or energy sourced from peatlands (Pin Koh 2007)</p> | <p>May reduce land available for smallholders in tropical peatlands (Jewitt et al 2014)</p> | <p>Will reduce supply of some materials sourced from peatlands (e.g palm oil, timber) (Murdiyarso et al. 2010)</p> | <p>Natural ecosystems are often source of medicines (UNEP, 2016)</p> | <p>Natural ecosystems often inspire learning (Turtle et al., 2015)</p> | <p>Natural environments support psychological wellbeing (Coldwell & Evans, 2018)</p> | <p>Natural environments support psychological wellbeing (Coldwell & Evans, 2018)</p> | <p>Retaining natural ecosystems can preserve genetic diversity (Ekins et al., 2003).</p> |

| | | | | | | | | | | | | | | | | | | | |
|-------------------------------|--|--|---|--|---|--|--|--|--|-----|-----|---|---|-----|-----|-----|--|---|---|
| | Enhanced weathering of minerals | N/A | N/A | N/A | See main text for mitigation potentials | Addition of basic minerals counteracts ocean acidification (Taylor et al., 2016) | N/A | May have negative effects on water quality (Atekwane et al. 2005) | Could improve soil quality (Rau & Caldiera 1999; Kantola et al 2017) | N/A | N/A | N/A | Can contribute to increase food production by replenishing plant available silicon, potassium and other plant nutrients (Beerling et al., 2018) | N/A | N/A | N/A | N/A | N/a | N/A |
| Carbon dioxide removal | Bioenergy and BECCS | Likely will reduce natural habitat with negative effects on biodiversity (Hof et al. 2018) | Would reduce natural pollinators due to decreased natural habitat if in competition (Keitt 2009). | The use of BECCS could reduce air pollution (SR15) | See main text for mitigation potentials | Mitigation potential (see main text) will reduce ocean acidification. | Will likely require water for plantations of fast growing trees and models show high risk of water scarcity if BECCS is deployed on widespread scale (Popp et al 2011; Smith et al. 2016; Hejazi et al., 2014) through both increases in water withdrawals (Hejazi et al., 2014; Bonsch et al., 2015) and changes in surface runoff (Cibin et al., 2015) | Bioenergy can affect freshwater quality via changes in nitrogen runoff from fertiliser application. However, the sign of the effect depends on what would have happened absent any bioenergy production, with some studies indicating improvements in water quality (Ng et al., 2010) and others showing declines (Sinha et al., 2019) | Will likely decrease soil quality if exotic fast growing trees used (Stoy et al. 2018) | N/A | N/A | BECCS and biofuels can contribute up to 300 EJ of primary energy by 2100 (Clarke et al., 2014). | BECCS will likely lead to significant trade-offs with food production (Smith et al 2016; Popp et al., 2017; Fujimori et al., in review) | N/A | N/A | N/A | BECCS would drive land use conversion and reduce opportunities for recreation/tourism. | BECCS would drive land use conversion and reduce culturally significant landscapes. | BECCS would drive land use conversion and reduce genetic diversity. |

1

2

Table SM6.10 Impacts on Nature’s Contributions to People of integrated response options based on value chain management

| <u>Integrated response options based on value chain management</u> | Habitat creation and maintenance | Pollination and dispersal of seeds and other propagules | Regulation of air quality | Regulation of climate | Regulation of ocean acidification | Regulation of freshwater quantity, flow and timing | Regulation of freshwater and coastal water quality | Formation, protection and decontamination of soils and sediments | Regulation of hazards and extreme events | Regulation of organisms detrimental to humans | Energy | Food and feed | Materials and assistance | Medicinal, biochemical and genetic resources | Learning and inspiration | Physical and psychological experiences | Supporting identities | Maintenance of options | |
|--|----------------------------------|---|---------------------------|-----------------------|---|--|--|--|--|---|--------|---------------|---|--|--------------------------|--|-----------------------|------------------------|-----|
| Demand management | Dietary change | Will lead to reduced expansion of agricultural lands, which can increase natural habitat (Tilman et al. 2001) | N/A | N/A | See main text on climate mitigation impacts | N/A | Will reduce water consumption if less water-intensive food/livestock needs to be produced (Tilman et al. 2001) | Reduced meat consumption will improve water quality (Stoll-Kleeman & O’Riordan 2015) | N/A | N/A | N/A | N/A | Will help increase global food supplies (Kastner et al. 2012) | N/A | N/A | N/A | N/A | N/A | N/A |

| | | | | | | | | | | | | | | | | | | | | |
|--------------------------|--|--|-----|--|---|-----|---|--|-----|-----|---|---|--|---|---|-----|-----|-----|-----|-----|
| | Reduced post-harvest losses | Will lead to reduced expansion of agricultural lands, which can increase natural habitat (Tilman et al. 2001) | N/A | N/A | See main text on climate mitigation impacts | N/A | Will reduce water consumption if less water-intensive food/livestock needs to be produced (Tilman et al. 2001) | N/A | N/A | N/A | Reducing postharvest losses will include measures to deal with pests, some of which could be biological (Wilson & Pusey 1985) | N/A | Will help increase global food supplies (Kastner et al. 2012) | N/A | N/A | N/A | N/A | N/A | N/A | |
| | Reduced food waste (consumer or retailer) | Improved storage and distribution reduces food waste and the need for compensatory intensification of agricultural areas thereby creating co-benefits for reduced land degradation (Stathers et al. 2013). | | | See main text on climate mitigation impacts | | Will reduce water consumption if less water-intensive food/livestock needs to be produced (Tilman et al. 2001) | Reduced food production will reduce N fertiliser use, improving water quality (Kibler et al. 2018) | N/A | N/A | N/A | N/A | Will help increase global food supplies (Kastner et al. 2012) | N/A | N/A | N/A | N/A | N/A | N/A | |
| | Material substitution | Material substitution increases demand for wood, which can lead to loss of habitat (Sathre & Gustavsson 2006). | | | See main text on climate mitigation impacts | N/A | N/A | N/A | N/A | N/A | N/A | N/A | Material substitution supplies building materials to replace concrete and other nonrenewables (Gustavsson & Sathre 2011) | N/A | N/A | N/A | N/A | N/A | N/A | |
| Supply management | Sustainable sourcing | Forest certification and other sustainable sourcing schemes can reduce habitat fragmentation as compared to conventional supply chains (Brown et al. 2001; Rueda et al. 2015)) | N/A | Forest certification improved air quality in Indonesia by 5% due to reduced incidence of fire (Miteva et al. 2015) | N/A | N/A | Forest certification has led to improved water flow due to decreased road construction for logging (Miteva et al. 2015) | Forest certification has improved riparian waterways and reduced chemical inputs in some schemes (Rueda et al. 2015) | N/A | N/A | N/A | Sustainable sourcing can supply energy like biomass (Sikkema et al. 2014) | Sustainable sourcing can supply food and other goods (G. Smith 2007) | Sustainable sourcing is increasingly important in timber imports (Ireland 2008) | Sustainable sourcing can supply medicinals (Pierce & Laird 2003). | N/A | N/A | N/A | N/A | N/A |
| | Management of supply chains | | N/A | Better management of supply chains may reduce energy use and air pollution in transport (Zhu et al. | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | Improved supply chains will help increase global food supplies (Hamprecht 2005). | Improved supply chains will help increase material supplies due to efficiency gains (Burritt & Schaltegger 2014). | N/A | N/A | N/A | N/A | N/A | |

| | | | | | | | | | | | | | | | | | | | |
|--|--|---|--|---|-----|--|--|-----|-----|-----|-----|-----|---|-----|-----|--|-----|--|---|
| | | | | 2018) | | | | | | | | | | | | | | | |
| Enhanced urban food systems | Urban gardening can improve habitat and biodiversity in cities (Orsini et al. 2014; Lin et al. 2015) | Urban beekeeping has been important in keeping pollinators alive (Gunnarsson & Federsel 2014) | Urban agriculture can increase vegetation cover and improve air quality in urban areas (Cameron et al. 2012; Lin et al. 2015). | See main text on climate mitigation impacts | N/A | Water access often a constraint on urban agriculture and can increase demands (De Bon et al 2010; Badami & Ramankutty 2015). | Urban agriculture can exacerbate urban water pollution problems (pesticide runoff, etc) (Pothukuchi & Kaufmann 1999) | N/A | N/A | N/A | N/A | N/A | Local urban food production is often more accessible to local populations and can increase food security (Eigenbrod & Gruda 2015) | N/A | N/A | Urban agriculture can be used for teaching and learning (Travaline & Hunold 2010). | N/A | Urban agriculture can promote cultural identities (Baker 2004) | Urban food can contribute to preserving local genetic diversity |
| Improved food processing and retail | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |
| Improved energy use in food systems | N/A | N/A | N/A | See main text on climate mitigation impacts | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A |

1

2

Table SM6.11 Impacts on Nature’s Contributions to People of integrated response options based on risk management

| <u>Integrated response options based on risk management</u> | Habitat creation and maintenance | Pollination and dispersal of seeds and other propagules | Regulation of air quality | Regulation of climate | Regulation of ocean acidification | Regulation of freshwater quantity, flow and timing | Regulation of freshwater and coastal water quality | Formation, protection and decontamination of soils and sediments | Regulation of hazards and extreme events | Regulation of organisms detrimental to humans | Energy | Food and feed | Materials and assistance | Medicinal, biochemical and genetic resources | Learning and inspiration | Physical and psychological experiences | Supporting identities | Maintenance of options | |
|---|--|--|---|---|-----------------------------------|---|---|---|--|---|--------|--|--------------------------|--|--------------------------|--|-----------------------|------------------------|-----|
| Management of urban sprawl | Reducing urban sprawl can help preserve natural habitat in periurban areas (Pataki et al 2011) | Reducing urban sprawl will help reduce loss of natural pollinators from habitat conversion (Cane 2005) | Urban sprawl is a major contributor to air pollution (Frumkin 2002) | See main text on climate mitigation impacts | | Managing urban sprawl can increase water availability (Pataki et al 2011) | Urban sprawl is associated with higher levels of water pollution due to loss of filtering vegetation and increasing impervious surfaces (Romero & Ordenes 2004; Tu et al 2007; Pataki et al 2011) | Likely to be beneficial for soils as soil sealing is major problem in urban areas (Scalenghe & Marsan 2009) | N/A | N/A | | Urban sprawl often competes with land for food production and can reduce overall yields (Chen 2007, Barbero-Sierra et al., 2013) | N/A | N/A | N/A | N/A | N/A | N/A | N/A |

| | | | | | | | | | | | | | | | | | | |
|-----------------------------------|---|---|-----|-----|-----|--|---|---|---|---|-----|--|--|--|---|-----|--|--|
| Livelihood diversification | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | Diversification is associated with increased access to income and additional food sources for the household (Pretty et al. 2003) | Diversification can increase access to materials (Smith et al. 2017) | N/A | N/A | N/A | N/A | N/A |
| Use of local seeds | Use of commercial seeds can contribute to habitat loss (Upreti & Upreti 2002) | Use of open pollinated seeds is beneficial for pollinators and creates political will to conserve them (Helicke 2015) | N/A | N/A | N/A | Local seeds often have lower water demands, as well as less use of pesticides that can contaminate water (Adhikari 2014) | Likely to contribute to less pollution as local seeds are usually grown organically (Adhikari 2014) | Likely to contribute to better soils as local seeds are usually grown organically (Adhikari 2014) | N/A | Local seeds often need less pesticides thereby reducing pest resistance (Adhikari 2014) | N/A | Local seeds can lead to more diverse and healthy food in areas with strong food sovereignty networks (Coomes et al. 2015; Bisht et al. 2018). However local seeds often are less productive than improved varieties. | | Many local seeds can have multiple functions, including medicinals (Hammer & Teklu 2008) | Passing on seed information is important cultural learning process (Coomes et al. 2015) | | Seeds associated with specific cultural identities for many (Coomes et al. 2015) | Food sovereignty movements have promoted saving of genetic diversity of crops through on-farm maintenance (Isakson 2009) |
| Disaster risk management | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | DRM helps people avoid extreme events and adapt to climate change (Mechler et al. 2014) | N/A | N/A | Famine early warning systems have been successful in Sahelian Africa to alert authorities to impending food shortages so that food acquisition and transportation from outside the region can begin, potentially helping millions of people (Genesio et al. 2011; Hillbruner and Moloney 2012) | | | | | | N/A |

| | | | | | | | | | | | | | | | | | | |
|--|--|---|---|-----|-----|-----|-----|---|--|-----|--|-----|--|--|-----|-----|-----|--|
| | | Commercial crop insurance often encourages habitat conversion; Wright and Wimberly (2013) found a 531,000 ha decline in grasslands in the Upper Midwest of the US 2006-2010 due to crop conversion driven by higher prices and access to insurance. | Crop insurance is likely to impact natural pollinators due to incentives for production (Horowitz & Lichtenberg 1993) | N/A | N/A | N/A | N/A | Likely to have negative effect as crop insurance encourages more pesticide use (Horowitz & Lichtenberg 1993). | One study found a 1% increase in farm receipts generated from subsidised farm programs (including crop insurance and others) increased soil erosion by 0.135 tons per acre (Goodwin and Smith 2003). | N/A | Crop insurance increases nitrogen use and leads to treating more acreage with both herbicides and insecticides (Horowitz & Lichtenberg 1993) | N/A | Crop insurance has generally lead to (modest) expansions in cultivated land area and increased food production (Claassen et al. 2011; Goodwin et al. 2004) | Insurance encourages monocropping leading to loss of genetic diversity for future (Glauber 2004) | N/A | N/A | N/A | Insurance encourages monocropping leading to loss of genetic diversity for future (Glauber 2004) |
|--|--|---|---|-----|-----|-----|-----|---|--|-----|--|-----|--|--|-----|-----|-----|--|

1
2
3
4

Table SM6.12 Impacts on the UN SDG of integrated response options based on land management

| <u>Integrated response options based on land management</u> | GOAL 1: No Poverty | GOAL 2: Zero Hunger | GOAL 3: Good Health and Well-being | GOAL 4: Quality Education | GOAL 5: Gender Equality | GOAL 6: Clean Water and Sanitation | GOAL 7: Affordable and Clean Energy | GOAL 8: Decent Work and Economic Growth | GOAL 9: Industry, Innovation and Infrastructure | GOAL 10: Reduced Inequality | GOAL 11: Sustainable Cities and Communities | GOAL 12: Responsible Consumption and Production | GOAL 13: Climate Action | GOAL 14: Life Below Water | GOAL 15: Life on Land | GOAL 16: Peace and Justice Strong Institutions | GOAL 17: Partnerships to achieve the Goal |
|---|--|--|--|---------------------------|--|---|-------------------------------------|--|---|---|--|---|--|---|--|--|---|
| Agriculture Increased food productivity | Increasing farm yields for smallholders contributes to poverty reduction (Irz et al 2001; Pretty et al 2003) | Increasing farm yields for smallholders reduces food insecurity (Irz et al 2001; Pretty et al 2003). | Increased food productivity leads to better health status (Rosegrant & Cline 2003; Dar & Gowda 2011) | N/A | Increased productivity can benefit female farmers, who make up 50% of agricultural labor in sub-Saharan Africa (Ross et al 2015) | Food productivity increases could impact water quality if increases in chemicals used, but evidence is mixed on sustainable intensification (Rockstrom et al 2009; Mueller et al 2012). | N/A | Increased agricultural production generally (Lal 2006) contributes to increased economic growth. | N/A | Increased agricultural production can contribute to reducing inequality among smallholders (Datt & Ravallion 1998). | Increased food production can increase urban food security (Ellis & Sumberg 1998). | N/A | See main text on climate mitigation and adaptation | Increased food productivity might be achieved through increased pesticide or fertiliser use, which causes runoff and dead zones in oceans (Beusen et al 2016) | See main text on desertification and degradation | N/A | Improved agricultural productivity generally correlates with increases in trade in agricultural goods (Fader et al. 2013) |

| | | | | | | | | | | | | | | | | | | |
|-------------------------------------|--|--|---|-----|---|--|--|---|--------------------------|---|-----|-----|---|--|-----|--|-----|-----|
| | | et al. 2012). | al. 2012). | | | | Improved livestock management can contribute to better water quality such as through manure management (Herrero & Thornton 2013) | | global South (Mack 1990) | | | | (de Wit et al 1995). | | | | | |
| Agro-forestry | Agroforestry can be usefully used for poverty reduction (Leakey & Simons 1997). | Agroforestry contributes to food productivity and reduces food insecurity (Mbow et al. 2014). | Agroforestry positively contributes to food productivity and nutritious diets (Haddad 2000) | N/A | Increased use of agroforestry can benefit female farmers as it requires low overhead, but land tenure issues must be paid attention to (Kiptot & Franzel 2012). | Agroforestry can be used to increase ecosystem services benefits, such as water quantity and quality (Jose 2009) | Agroforestry could increase biomass for energy (Mbow et al. 2014) | Agroforestry and other forms of employment in forest management make major contributions to global GDP (Pimental et al 1997). | N/A | Agroforestry promotion can contribute to reducing inequality among smallholders (Lefmeister et al 2018). | N/A | | Agroforestry contributes to sustainable production goals (Mbow et al 2014). | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| Agricultural diversification | Agricultural diversification is associated with increased welfare and incomes and decreased levels of poverty in several country studies (Arslan et al. 2018; Asfaw et al. 2018; Weinberger & Lumpkin 2007). | Diversification is associated with increased access to income and additional food sources for the farming household (Pretty et al. 2003; Ebert 2014). Diversification can also reduce the risk of crop pathogens spreading across landscapes (Lin 2011). | More diversified agriculture leads to diversified diets which have better health outcomes (Block & Webb 2001; Ebert 2014; Kadiyala et al 2014) particularly for women and children (Pretty et al. 2003) | N/A | N/A | N/A | N/A | Agricultural diversification can lead to economic growth (Rahman 2009; Pingali & Rosegrant 1995). It allows farmers to choose a strategy that both increases resilience and provides economic benefits, including functional biodiversity at multiple spatial and/or temporal scales, through practices developed via | N/A | Increased agricultural diversification can contribute to reducing inequality among smallholders (Makate et al 2016), although there is mixed evidence of inequality also increasing in commercialised systems (Pingali & Rosegrant 1995; Weinberger & Lumpkin 2007) | N/A | N/A | N/A | See main text on desertification and degradation | N/A | N/A | N/A | |

| | | | | | | | | | | | | | | | | | |
|---|--|--|--|-----|---|---|-----|--|-----|--|--|--|--|--|--|--|-----|
| | | | | | | | | traditional and/or agroecological scientific knowledge (Lin 2011 ; Kremen et al. 2012). | | | | | | | | | |
| Avoidance of conversion of grassland to cropland | May reduce land available for cropping or livestock for poorer farmers ; some grassland restoration programs in China have been detrimental to poor pastoralists (Foggin 2008) | Can affect food security when competition for land occurs (O'Mara 2012) | N/A | N/A | N/A | Retaining grasslands contributes to better water retention and improved quality (Scanlon et al 2007). | N/A | Reduced cropland expansion may decrease GDP (Lewandrowski et al 1999) | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| Integrated water management | Green water harvesting contributes to alleviate poverty in Sub-Saharan Africa (Rockström and Falkenmark 2015). Improving water irrigation (Rengasamy 2006), improving rainfed agriculture (integrating soil and water management, rainfall infiltration and water harvesting, provides a large co-benefit to delivery of food security and poverty reduction (UNCTAD 2011) | Integrated, efficient, equitable and sustainable water resource management (as water for agroecosystem) plays importance for food production and benefits to people (Lloyd et al. 2013). | Water is a finite and irreplaceable resource that is fundamental to human well-being. It is only renewable if well managed. Integrated water management is vital option for reducing the global burden of disease and improving the health, welfare and productivity of populations. Today, more than 1.7 billion people live in river basins where depletion through use exceeds natural recharge, a trend that | N/A | Involving both women and men in integrated water resources initiatives can increase project effectiveness and efficiency (Green & Baden 1995) | Water resource management is intended to solve watershed problems on a sustainable basis, and these problems can be categorised into lack of water (quantity), deterioration in water quality, ecological effects, poor public participation, and low output economic value for investment in watershed-related activities (Lee et al. 2018). Integrated water management, increase | N/A | Water is at the core of sustainable development and is critical for socio-economic development, healthy ecosystems and for human survival itself. Integrated water management can play a key enabling role in strengthening the resilience of social, economic and environmental systems in the light of rapid and unpredictable changes | N/A | IWM can increase access of industry to water for economic growth (Rahman & Varis 2005) | Water is a limiting factor in urban growth and IWM can help improve access to urban water supplies (Bao & Fang 2012) | Poor sectoral coordination and institutional fragmentation have triggered an unsustainable use of resources and threatened the long-term sustainability of food, water, and energy security (Rassul 2016). | See main text on climate mitigation and adaptation | IWM on land is likely to improve water quality runoff into oceans (Agboola & Braimah 2009) | See main text on desertification and degradation | Integrated water management, increase water-use efficiency across all sectors and ensure sustainable withdrawals and supply of freshwater to address water scarcity, and substantially reduce the number of people suffering from water scarcity (UN Water, 2015). | |

| | | | | | | | | | | | | | | | | | | |
|----------|--|---|--|---|-----|--|--|--|---|--|-----|--|--|--|-----|--|---|---|
| | | | | will see two-thirds of the world's population living in water-stressed countries by 2025 (UNWater 2015) | | | water-use efficiency across all sectors and ensure sustainable withdrawals and supply of freshwater to address water scarcity, and substantially reduce the number of people suffering from water scarcity (UNWater 2015). | | (UN Water, 2015). | | | | | | | | | |
| Forestry | Forest management and forest restoration | May contribute to poverty reduction if conditions are right (Blomley & Ramadhani 2006; Donovan et al 2006), but conflicting data, as it may also favor large landowners who are less poor (Rametsteiner and Simula 2003). | Forest expansion can affect crop production when competition for land occurs (Angelsen 2010). An increase in global forest area can lead to increases in food prices through increasing land competition (Calvin et al. 2014b; Kreidenweis et al. 2016c; Reilly et al. 2012b; Smith et al. 2013a; Wise et al. 2009b) | N/A | N/A | Women face challenges in sustainable forest management (Mwangi et al 2011), but N/A how SFM affects gender equity. | Forests tend to maintain water quality by reducing runoff and trapping sediments and nutrients (Idris Medugu et al. 2010c; Salvati et al. 2014a). Due to evapotranspiration, trees recharge atmospheric moisture, contributing to rainfall locally and in distant location, and trees' microbial flora and biogenic volatile organic compounds can directly promote rainfall (Arneeth et al. 2010). Trees enhance soil infiltration and, under | SFM may increase availability of biomass for energy (Kraxner et al. 2013; Sikkema et al. 2013) | Forest management often require employment for active replanting, etc. (Ros-Tonen et al 2008) | Forestry supplies wood for industrial use (Gustavsson & Sathre 2011) | N/A | Community forest management can contribute to stronger communities (Padgee et al 2006) | Improved forest management contributes to sustainable production goals, e.g. thru certification of timber (Ramets teiner and Simula 2003). | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | Sustainable forest management often requires collective action institutions (Ros-Tonen et al 2008). | Sustainable forest management can contribute to increases in demand for wood products (e.g. certification) (McDonald & Lane 2004) |

| | | | | | | | | | | | | | | | | | | |
|--|--|--|--|---|-----|-----|---|--|---|-----|-----|-----|-----|--|-----|--|-----|-----|
| | | participants | | | | | microbial flora and biogenic volatile organic compounds can directly promote rainfall (Armeth et al. 2010). Trees enhance soil infiltration and, under suitable conditions, improve groundwater recharge (Calder 2005; Ellison et al. 2017a; Neary et al. 2009b). | | | | | | | | | | | |
| | | May contribute to poverty reduction but conflicting data (Tschakert 2007). Many projects for reforestation may have some small impacts on poor households, while others actually increased poverty due to land losses or lack of economic impacts (Jindal et al 2008). | Forest expansion can affect crop production when competition for land occurs (Angelsen 2010). An increase in global forest area can lead to increases in food prices through increasing land competition (Calvin et al. 2014b; Kreidenweis et al. 2016c; Reilly et al. 2012b; Smith et al. 2013a; Wise et al. 2009b) | Reforestation can enhance human well-being by microclimatic regulation for protecting people from heat stresses (Locatelli et al. 2015c) and generally improve the cultural and recreational value of ecosystems (Knoke et al. 2014). Trends of forest resources of nations are found to positively correlate with UNDP Human Development Index (Kauppi et al. 2018). | N/A | N/A | Particular activities associated with forest landscape restoration, such as mixed planting, assisted natural regeneration, and reducing impact of disturbances (e.g. prescribed burning) have positive implications for fresh water supply (Ciccarese et al. 2012; Suding et al. 2015). | Reforestation can increase availability of biomass for energy use (Swischer 1994). | Reforestation often require employment for active replanting, etc. (Jindal et al. 2008) | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| | | Although some have argued that afforestation can be a tool for poverty | Future needs for food production are a constraint for large-scale afforestation | Afforestation can enhance human well-being by microclimatic | N/A | N/A | Afforestation using some exotic species can upset the | Afforestation may increase availability of biomass for energy use (Obersteiner | Afforestation often requires employment for active | N/A | N/A | N/A | N/A | See main text on climate mitigation and | N/A | See main text on desertification and degradation | N/A | N/A |

| | | | | | | | | | | | | | | | | | | |
|-----------------|---------------------------------------|--|---|---|-----|---|---|-------------|---|-----|--|-----|-----------------------------------|---|---|--|-----|-----|
| | | reduction (Holden et al 2003), afforestation can compete with land available for cropping and poor farmers often do not benefit from afforestation projects (McElwee 2009) | plans (Locatelli et al. 2015c). Global food crop demand is expected by 50%-97% between 2005 and 2050 (Valin et al. 2014). Future carbon prices will facilitate deployment of afforestation projects at expenses of food availability (adverse side-effect), but more liberalised trade in agricultural commodities could buffer food price increases following afforestation in tropical regions (Kreidenweis et al. 2016c) | ic regulation for protecting people from heat stresses (Locatelli et al. 2015c) and generally improve the cultural and recreational value of ecosystems (Knoke et al. 2014). Trends of forest resources of nations are found to positively correlate with UNDP Human Development Index (Kauppi et al. 2018) | | | balance of evapotranspiration regimes, with negative impacts on water availability particularly in arid regions (Ellison et al. 2017a; Locatelli et al. 2015; Trabucco et al. 2008). Afforestation in arid and semiarid regions using species that have evapotranspiration rates exceeding the regional precipitation may aggravate the groundwater decline (Locatelli et al. 2015a; Lu et al. 2016). Changes in runoff affect water supply but can also contribute to changes in flood risks, and irrigation of forest plantations can increase water consumption (Sterling et al. 2013) | et al 2006) | replanting . etc. (Mather & Murray 1987). | | | | | adaptati on | | | | |
| Soil management | Increased soil organic carbon content | Can increase yields for smallholders, which can contribute to poverty | Lal (2006b) notes that "Food-grain production in developing countries can | There is evidence that increasing soil organic carbon | N/A | Gender impacts use of soil organic matter practices | Soil organic matter is known to increase water | N/A | Increased agricultural production generally | N/A | Increased agricultural production can contribute | N/A | Improved conservation agriculture | See main text on climate mitigation and | Rivers transport dissolved organic matter to oceans | See main text on desertification and degradation | N/A | N/A |

| | | | | | | | | | | | | | | | | | | |
|--|----------------------------------|--|--|---|-----|---|--|-----|---|-----|--|---|--|--|--|--|-----|--|
| | | reduction, but because adoption often depends on exogenous factors these need to be taken into consideration (Wollni et al 2010; Kassie et al 2013). | be increased by 24-39 (32+-11) million Mgy-1 through improving soil quality by increasing the SOC pool and reversing degradation processes". | could be effective in reducing the prevalence of disease-causing helminths (Lal 2016; Wall et al. 2015). Also indirectly contributes to food productivity which may have impact on diets. | | (Quansah et al 2001) but N/A how the relationship works in reverse. | filtration and protects water quality (Lehmann & Kleber 2015) | | (Lal 2006c) contributes to increased economic growth. | | to reducing inequality among smallholders (Datt & Ravallion 1998). | | contributes to sustainable production goals (Hobbs et al. 2008). | adaptation | (Hedges et al 1997), but unclear if improved SOM will decrease this and by how much. | | | |
| | | | | | | | Various researchers showed a relationship between impact of soil erosion and degradation on water quality indicating the source of pollutant as anthropogenic and industrial activities. in China (Issaka & Ashraf 2017). Managing soil erosion improves water quality (Pimentel et al 1995) | | | | | Particulate matter pollution, a main consequence of wind erosion, imposes severe adverse impacts on materials, structures and climate which directly affect the sustainability of urban cities (Al-Thani et al. 2018) | | See main text on climate mitigation and adaptation | | See main text on desertification and degradation | | |
| | Reduced soil erosion | Can increase yields for smallholders and contributes to poverty reduction (Ananda & Herath 2003) | Contributes to agricultural productivity and reduces food insecurity (Pimentel et al. 1995; Shiferaw & Holden 1999). | Contributes to food productivity and improves farmer health (Pimentel et al. 1995; Shiferaw & Holden 1999). | N/A | N/A | | N/A | N/A | N/A | N/A | | N/A | | N/A | | N/A | |
| | Reduced soil salinisation | Salinisation can impoverish farmers (Duraiappah 1998) therefore preventing or reversing can increase yields for smallholders and contributes to poverty reduction. | Reversing degradation contributes to food productivity and reduces food insecurity (Pimentel et al. 1995; Shiferaw & Holden 1999). | Salinisation is known to have human health impacts: wind-borne dust and respiratory health; altered ecology of mosquito-borne diseases; and mental health consequences (Jardine | N/A | N/A | Management of soil salinity improves water quality and quantity (Kotb et al. 2000; Zalidis et al 2002) | | | | | | See main text on climate mitigation and adaptation | | See main text on desertification and degradation | | | |

| | | | | | | | | | | | | | | | | | | |
|-----------------------------------|---|--|--|---|-----|-----|--|-----|-----|---------------|-----|-----------------------------|-----|--|------------------------|--|-----|-----|
| | | | | et al 2007) | | | | | | | | | | | | | | |
| | Reduced soil compaction | Soil compaction and other forms of degradation can impoverish farmers (Scherr 2000); prevention of compaction thus contributes to poverty reduction. | Compactions reduces agricultural productivity and thus contributes to food insecurity (Nawaz et al 2013) | Soil compaction has human health consequences as it contributes to runoff of water and pollutants into surface and groundwater (Soane and van Ouwerkerk 1994) | N/A | N/A | Management of soil compaction improves water quality and quantity (Soane and van Ouwerkerk 1994; Zalidis et al 2002) | N/A | N/A | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| | Biochar addition to soil | Land to produce biochar may reduce land available for smallholders, and it tends to be unaffordable for poor farmers; as of yet, few biochar projects have shown poverty reduction benefits (Leach et al 2012) | Could potentially affect crop production if competition for land occurs (Ennis et al 2012) | N/A | N/A | N/A | Biochar improves soil water filtration and retention (Spokas et al 2011) | N/A | N/A | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| Other ecosystem management | Fire management | N/A | N/A | Fire management reduces health risks from particulates (Bowman & Johnston 2005). | N/A | N/A | Fires affect water quality and flow due to erosion exposure (Townsend & Douglas 2000). | N/A | N/A | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| | Reduced landslides and natural hazards | Landslides can increase vulnerability to poverty (Msilimba 2010), therefore management will reduce risks to the poor | Landslides are one of the natural disasters that have impacts on food security (de Haen & Henrich 2007) | Managing landslides reduces health risks (Haines et al 2006) | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| | Reduced pollution | N/A | N/A | Reducing acid | N/A | N/A | Pollution increases | N/A | N/A | Management of | N/A | Management of pollution can | N/A | See main | Reduction in pollution | See main text on | N/A | N/A |

| | | | | | | | | | | | | | | | | | | |
|--|---|---|--|---|-----|-----|---|--|--|--|-----|---|-----|--|--|--|-----|-----|
| | including acidification | | | deposition reduces health risks, including respiratory illnesses and increased morbidity (Lübker-Alcamo & Krzyzanowski 1995; Larssen et al 1999) | | | acidity of surface water, with likely ecological effects (Larssen et al 1999) | | | pollution can increase demand for new technologies (Popp 2006). | | reduce exposure to health risks in urban areas (Bartone 1991) | | text on climate mitigation and adaptation | can improve water quality running to oceans (Doney et al 2007). | desertification and degradation | | |
| | Management of invasive species / encroachment | Invasive species removal policies have been beneficial to the poor (van Wilgen & Wannenburgh 2016) | IAS can compete with crops and reduce yields by billions of dollars annually (Pejchar & Mooney 2009) | IAS have strong negative effects on human well-being (Pejchar & Mooney 2009) | N/A | N/A | IAS like the golden apple snail/zebra mussel have damaged aquatic ecosystems (Pejchar & Mooney 2009) | N/A | IAS removal policies can increase employment due to need for labor (van Wilgen & Wannenburgh 2016) | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| | Restoration and avoided conversion of coastal wetlands | Impacts on poverty are mixed (Kumar et al 2011). May reduce land available for cropping, and poor design can impoverish people (Ingram et al 2006; Mangora 2011). Can also decrease vulnerability to coastal storms, however (Jones et al. 2012; Feagin et al 2010) | Mixed evidence: can affect agriculture/fisheries production when competition for land occurs, or could increase food production when ecosystems are restored (Crooks et al 2011) | Wetlands contribute to local well-being (Crooks et al 2011), and restoration generally improve the cultural and recreational value of ecosystems (Knoke et al. 2014). | N/A | N/A | Wetlands store freshwater and enhance water quality (Bobbink et al 2006) | N/A | Restoration projects often require employment for active replanting, etc. (Crooks et al. 2011). | Protecting coastal wetlands may reduce infrastructure projects in coastal areas (e.g. sea dikes, etc.) (Jones et al. 2012) | N/A | N/A | N/A | See main text on climate mitigation and adaptation | Restoration of coastal wetlands can play a large role in providing habitat for marine fish species (Bobbink et al 2006; Hale et al 2009) | See main text on desertification and degradation | N/A | N/A |
| | Restoration and avoided conversion of peatlands | May reduce land available for smallholders in tropical peatlands (Jewitt et al 2014) | Can affect crop production when competition for land occurs, although much use of peatlands in tropics is for palm oil, not food (Sellamuttu et al 2011) | | N/A | N/A | Peatland restoration will improve water quality as they play important roles in water retention and drainage (Johnston 1991). | Peatlands in tropics are often used for biofuels and palm oil, so may reduce the availability of these (Danielsen et al 2008). | Reduced peatland exploitation may decrease GDP in Southeast Asia (Koh et al 2011) | N/A | N/A | N/A | N/A | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |

| | | | | | | | | | | | | | | | | | | |
|--|--|---|--|-----|-----|--|---|-----|---|-----|-----|-----|---|--|--|-----|--|--|
| | | | <p>Biodiversity, and its management, is crucial for improving sustainable and diversified diets (Global Panel on Agriculture and Food Systems for Nutrition 2016). Indirectly, the loss of pollinators (due to combined causes, including the loss of habitats and flowering species) would contribute to 1.42 million additional deaths per year from non-communicable and malnutrition-related diseases, and 27.0 million lost disability-adjusted life-years (DALYs) per year (Smith et al. 2015). However, at the same time, some options to preserve biodiversity, like protected areas, may potentially conflict with food production by local communities (Molotoks et al. 2017).</p> | | | | | | | | | | | | | | | |
| | Biodiversity conservation | There is mixed evidence on the impacts of biodiversity conservation measures on poverty | Biodiversity, and its management, is crucial for improving sustainable and diversified diets (Global Panel on Agriculture and Food Systems for Nutrition 2016). | N/A | N/A | 33 out of 105 of the largest urban areas worldwide rely on biodiversity conservation measures such as protected areas for some, or all, of their drinking water (Secretariat of the Convention on Biological Diversity 2008) | Some biodiversity conservation measures might increase access to biomass supplies (Erb et al. 2012) | | | | | | Biodiversity conservation measures like protected areas can increase ocean biodiversity (Selig et al. 2014) | Indigenous peoples' roles in biodiversity conservation can increase institutions and conflict resolution (Garnett et al. 2018) | Indigenous peoples commonly link forest landscapes and biodiversity to tribal identities, association with place, kinship ties, customs and protocols, stories, and songs (Gould 2014; Lyver et al. 2017a, b). | | | |
| | Enhanced weathering of minerals | N/A | N/A | N/A | N/A | Mineral weathering can affect the chemical composition of soil and surface waters (Katz | N/A | N/A | Will require development of new technologies (Schuiling and Krijgsman | N/A | N/A | N/A | See main text on climate mitigation and adaptation | See main text on desertification and degradation | N/A | N/A | | |

| | | | | | | | | | | | | | | | | | | |
|-----|---------------------|---|---|--|------------------------------------|------------------------------------|---|--|---|--|------------------------------------|------------------------------------|--|---|--|---|------------------------------------|------------------------------------|
| | | | | | | | 1989) | | | 2006) | | | | | | | | |
| CDR | Bioenergy and BECCS | Bioenergy production could create jobs in agriculture, but could also compete for land with alternative uses. Therefore, bioenergy could have positive or negative effects on poverty rates among smallholders, among other social effects (IPCC 2018). | Biofuel plantations may lead to decreased food security through competition for land (Locatelli et al. 2015c). BECCS will likely lead to significant trade-offs with food production (Popp et al. 2011c; Smith et al. 2016b). | BECCS could have positive effects through improvements in air and water quality (IPCC 2018), but BECCS could have negative effects on health and wellbeing through impacts on food systems (Burns and Nicholson 2017). Additionally, there is a non-negligible risk of leakage of CO2 (IPCC 2018). | No direct interaction (IPCC 2018). | No direct interaction (IPCC 2018). | Will likely require water for plantations of fast growing trees and models show high risk of water scarcity if BECCS is deployed on widespread scale (IPCC 2018). | BECCS and biofuels can contribute up to 300 EJ of primary energy by 2100 (cross-chapter box 7 on bioenergy); bioenergy can provide clean, affordable energy (IPCC 2018). | Access to clean, affordable energy will help economic growth (IPCC 2018). | BECCS will require development of new technologies (Smith et al. 2016c). | No direct interaction (IPCC 2018). | No direct interaction (IPCC 2018). | Switching to bioenergy reduces depletion of natural resources (IPCC 2018). | See main text on climate mitigation and adaptation. | Reductions in carbon emissions will reduce ocean acidification. See main text on climate mitigation. | See main text on desertification and degradation. | No direct interaction (IPCC 2018). | No direct interaction (IPCC 2018). |

1
2
3
4

Table SM6.13 Impacts on the UN SDG of integrated response options based on value chain interventions

| Integrated response options based on value chain management | GOAL 1: No Poverty | GOAL 2: Zero Hunger | GOAL 3: Good Health and Well-being | GOAL 4: Quality Education | GOAL 5: Gender Equality | GOAL 6: Clean Water and Sanitation | GOAL 7: Affordable and Clean Energy | GOAL 8: Decent Work and Economic Growth | GOAL 9: Industry, Innovation and Infrastructure | GOAL 10: Reduced Inequality | GOAL 11: Sustainable Cities and Communities | GOAL 12: Responsible Consumption and Production | GOAL 13: Climate Action | GOAL 14: Life Below Water | GOAL 15: Life on Land | GOAL 16: Peace and Justice Strong Institutions | GOAL 17: Partnerships to achieve the Goal |
|---|--|---|---|-----------------------------------|-----------------------------------|--|--|--|---|---|--|---|--|--|--|--|---|
| Demand management Dietary change | Reduced meat consumption can free up land for other activities to reduce poverty (Röös et al. 2017; Stoll-Kleemann and O’Riordan 2015). However, reduced demand for livestock will have negative effect on pastoralists and could suppress | High-meat diets in developed countries may limit improvement in food security in developing countries (Rosegrant et al. 1999); dietary change can contribute to | Overnutrition contributes to worse health outcomes, including diabetes and obesity (Tilman and Clark 2014a; McMichael et al. 2007). Dietary change away from meat consumption | No direct interaction (IPCC 2018) | No direct interaction (IPCC 2018) | Reduced meat consumption will reduce water consumption. (Muller et al. 2017b) found that lower impact agriculture could be practiced if dietary change and | Dietary shifts away from meat to fish/fruits/vegetables increases energy use in the US by over 30% (Tom et al. 2016) | Health costs of meat-heavy diets add to health care costs and reduce GDP (Popkin 2008) | N/A | There are currently large discrepancies in diets between developed and developing nations (Sans & Combris 2015). Dietary change will reduce food inequality by reducing meat overconsumption in Western countries and | Dietary change is most needed in urbanised, industrialised countries and can help contribute to demand for locally grown fruits and vegetables (Tom et al. 2016) | A dietary shift away from meat can contribute to sustainable consumption by reducing greenhouse gas emissions and reducing cropland and pasture requirements (Stehfest et al. 2009; Bajželj | See main text on climate mitigation and adaptation | Dietary change away from meat might put increased pressure on fish stocks (Vranken et al. 2014; Mathijs 2015). Overall reduced emissions | See main text on desertification and degradation | N/A | N/A |

| | | | | | | | | | | | | | | | | | | |
|--|--|--|--|---|--|--|---|--|---|--|--|--|---|--|---|--|-----|---|
| | | demand for other inputs (grains) that would affect poor farmers (Garnett 2011; IPCC SR15) | food security goals (Godfray et al. 2010a; Bajželj et al. 2014) | has major health benefits, including reduced heart disease and mortality (Popkin 2008; Friel et al. 2008). Dietary change could contribute to 5.1 million avoided deaths per year (Springmann et al. 2016) | | | waste reduction were implemented, leading to lower GHG emissions, lower rates of deforestation, and decreases in use of fertiliser (nitrogen and phosphorus), pesticides, water and energy. However, Tom et al. (2016) found water footprints of fruit/veg dietary shift in the US to increase by 16% | | | | free up some cereals for consumption in poorer diets (Rosegrant et al. 1999) | | et al. 2014). | | would decrease rate of ocean acidification (Doney et al. 2009) | | | |
| | Reduced post-harvest losses | Reducing food losses from storage and distribution operation can increase economic well-being without additional investment in production activities (Bradford et al. 2018; Temba et al. 2016) | Reducing food losses increases food availability, nutrition, and lower prices (Sheahan and Barrett 2017b; Abass et al. 2014; Affognon et al. 2015) | Improved storage enhances food quality and can reduce mycotoxin intake (Bradford et al. 2018; Temba et al. 2016; Stathers et al. 2013; Tirado et al. 2010) especially in humid climates (Bradford et al. 2018). The perishability and safety of fresh foods are highly susceptible to temperature increase (Bisbis et al. 2018; Ingram et al. 2016a). | Reduced losses can increase income that could be spent on education, but no data available | Postharvest losses do have a gender dimension (Kaminski and Christiaensen 2014), but unclear if reducing losses will contribute to gender equality (Rugumamu 2009) | Kummu et al. (2012a) reported that 24% of global freshwater use and 23% of global fertiliser use is attributed to food losses. Reduced post harvest losses can decrease need for additional agricultural production and irrigation. | Reduced losses would reduce energy demands in production; 2030 +- 160 trillion BTU of energy were embedded in wasted food in 2007 in the US (Cuéllar and Webber 2010) | In East and Southern Africa, postharvest loss for six major cereals was USD1.6 billion or 15% of total production value; reducing losses would thus boost GDP substantially in developing countries with PHL (Hodges et al. 2011) | Reducing PHL can involve improving infrastructure for farmers and marketers (Parfitt et al. 2010) | Poorer households tend to experience more PHL, and thus reducing PHL can contribute to reducing inequality among farmers (Hodges et al. 2011). | N/A | Reducing PHL contributes to sustainable production goals (Parfitt et al. 2010) | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | Post harvest losses contribute to higher food prices and constraints on trade (Tefera 2012) |
| | Reduced food waste (consumer or retailer) | Food waste tends to rise as incomes rise (Parfitt et al. 2010; Liu et al. 2013), so it is not clear what the relationship to poverty is. Could be potentially beneficial as it would free up | People who are already food insecure tend not to waste food (Nahman et al. 2012). Reduced food waste would | Food waste can increase with healthier diets (Parizeau et al. 2015). Health and safety standards can restrict some approaches to reducing food | N/A | Reducing food waste within households often falls to women (Stefan et al. 2013) and can increase their labor workload | Kummu et al. (2012a) reported that 24% of global freshwater and 23% of global fertiliser is used in the production of | Reduced losses would reduce energy demands in production; 2030 +- 160 trillion BTU of energy were embedded in wasted food in 2007 in the US (Cuéllar and Webber 2010). Food waste can be a | Waste generation has grown faster than GDP in recent years (Thogerson 1996). Households in the UK throw out USD745 | Food waste could be an important source of needed chemicals for industrial development in resource constrained countries (Lin et | Wealthier households tend to waste more food (Parfitt et al. 2010), but unclear how reducing waste may contribute to reducing inequality. | There have been large increases in the throughput of materials such as the food-waste stream, import and solid-waste | Post-consumer food waste in industrialised countries (222 million ton) is almost as high as the total net production in sub-Saharan Africa (230 | See main text on climate mitigation and adaptation | Reducing food waste may be related to food packaging, which is a major source of ocean pollution, | See main text on desertification and degradation | N/A | Food waste can contribute to higher food prices and constraints on trade (Tefera 2012) |

| | | | | | | | | | | | | | | | | | | |
|--------------------------|------------------------------|---|---|---|--|--|--|--|---|--|--|--|---|--|---|--|-----|---|
| | | money to spend on other activities (Dorward 2012). Redistribution of food surplus to the poor could also have impacts on poverty (Papargyropoulou et al. 2014) | increase the supply of food (FAO 2011; Smith 2013), but it is unclear if this would benefit those who are food insecure in developing countries (Hertel and Baldos 2016). | waste (Halloran et al. 2014). Changes in packaging to reduce waste might have negative health impacts (e.g. increased contamination) (Claudio 2012) | | (Hebrok and Boks 2017). Women also generate more food waste and could be a site for intervention (Thyberg and Tonjes 2016) | food losses, so reduction in food waste could provide significant co-benefits for freshwater provision and on nutrient cycling (Kummu et al. 2012). Muller et al. (2017b) found that lower impact agriculture could be practiced if dietary change and waste reduction were implemented, leading to lower GHG emissions, lower rates of deforestation, and decreases in use of fertiliser (nitrogen and phosphorus), pesticides, water and energy. | sustainable source of biofuel (Uçkun Kiran et al. 2014) | of food and drink each year as food waste; South Africans throw out USD 7billion worth of food per year (Nahman and de Lange 2013). Reductions of postconsumer waste would increase household income (Hodges et al. 2011) | al. 2013) | | accumulation in urban areas (Grimm et al. 2008). Reducing compostable food waste reduces need for landfills (Smit and Nasr 1992; Zaman and Lehmann 2011) | million ton). (FAO 2011), thereby reducing waste contributes to sustainable consumption. | | but relationship is not known (Hornweg et al 2013) | | | |
| | Material substitution | N/A | Could increase demand for wood and compete with land for agriculture, but no evidence of this yet. | N/A | N/A | N/A | If water is used efficiently in production of wood, likely to be positive impact over cement production (Gustavsson and Sathre 2011) | Concrete frames require 60-80% more energy than wood (Börjesson and Gustavsson 2000). Material substitution can reduce embodied energy of buildings construction by up to 20% (Thormark 2006; Upton et al. 2008) | The relationship between material substitution and GDP growth is unclear (Moore et al. 1996) | Material substitution may reduce need for industrial production of cement etc. (Petersen and Solberg 2005) | N/A | Changing materials for urban construction can reduce cities' ecological footprint (Zaman and Lehmann 2013) | Material substitution is a form of sustainable production/consumption which replaces cement and other energy-intensive materials with wood (Fiksel 2006) | See main text on climate mitigation and adaptation | Overall reduced emissions rate of ocean acidification (Doney et al. 2009) | See main text on desertification and degradation | N/A | N/A |
| Supply management | Sustainable sourcing | Value adding has been promoted as a successful poverty reduction strategy in many countries (Lundy et al. 2002; Whitfield 2012; Swanson 2006). Volatility of food supply and food price spikes in 2007 increased the number of people | Poor farmers can benefit from value-adding and new markets (Bamman 2007) and may help to improve food security by increasing its economic performance | Value-chains can help increase the nutritional status of food reaching consumers (Fan et al. 2012) | Value-adding can increase income that could be spent on education, but no data available | Women are highly employed in value-added agriculture in many developing countries, but do not always gain substantive benefits (Dolan and Sorby 2003). | Value-added products might require additional water use (Guan and Hubacek 2007), but depends on context. | N/A | Value-adding and export diversification generates additional employment and expands GDP in developing countries in particular (Newfarmer et al. 2009) | Value adding can create incentives to improve infrastructure in processing (Delgado 2010). Expanding value chains can incorporate new sources of food producers into industrial systems of | Value-adding can be an important component of additional employment for poorer areas, and can contribute to reductions in overall inequality. However, data shows high-value agriculture | Value-adding can increase incentives to keep peri-urban agriculture, but faces threats from rising land prices in urban areas (Midmore and Jansen | Value-adding in agriculture (e.g. fair trade, organic) can be an important source of sustainable consumption and production (de Haen and Réquillart 2014) | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | Value-adding has a strong relationship to expanding trade in developing countries in particular (Newfarmer et al. 2009) |

| | | | | | | | | | | | | | | | | | |
|------------------------------------|--|---|---|---|--|--|--|---|---|---|---|---|--|-----|--|-----|---|
| | | under the poverty line by between 100 million people (Ivanic and Martin 2008) to 450 million people (Brinkman et al. 2009), and caused welfare losses of 3% or more for poor households in many countries (Zezza et al. 2009). | and revenues for local farmers (Reidsma et al. 2010). However, much value-adding is captured upstream, not by poor producers (McMichael and Schneider 2011b). Food prices strongly affect food security (Lewis and Wiham 2012; Regmi and Meade 2013; Fujimori et al. 2018a), and policies to decrease volatility will likely have strong impacts on food security (Timmer 2009; Torlesse et al. 2003b; Raleigh et al. 2015b). | | | Value-chains that target women could increase gender equity, but data is scarce (Gengenbach et al. 2018) | | | distribution (Bloom and Hinrichs 2011) | is not always a pathway toward enhanced welfare (Dolan and Sorby 2003), and much value-adding is captured not by smallholders but higher up the chain (Neilson 2007) | 2003) | | | | | | |
| Management of supply chains | Reducing food transport costs generally helps poor farmers (Altman et al. 2009). More than USD200 million is generated in fresh fruit and veg trade between Kenya and the UK; much has contributed to poverty reduction and better transport could increase the amount generated (MacGregor and Vorley 2006; Muriithi and Matz 2015). Volatility of food supply and food price spikes in 2007 increased the number of people under the poverty line by between 100 million people (Ivanic and Martin | Improving efficiency can reduce food waste and health risks associated with poor storage management practices (James and Bradford et al. 2018; Temba et al. 2016; Stathers et al. 2013; Tirado et al. 2010). There is some limited evidence that improved transport on-farm increases | Access to quality food is a major contributor to whether a diet is healthy or not (Neff et al. 2009). Increased distribution and access of packaged foods however can decrease health outcomes (Galal et al. 2010; Monteiro et al. 2011) | Reduction in staple food price costs to consumers in Bangladesh from food stability policies saved rural households USD887 million total (Torlesse et al. 2003b), but N/A if this increased spending on education in households | Women and girls are often the most effected ones in households when there are food shortages (Kerr 2005; Hadley et al. 2008) | Food imports can contribute to water scarcity through "embodied" or "virtual" water accounting (Yang and Zehnder 2002; Guan and Hubacek 2007; Hanjra and Qureshi 2010; Jiang 2009) | Food supply chains and flows have adverse effects due to reliance on non-renewable energy (Kurian 2017; Scott 2017). Shifts to biofuels can destabilise food supplies (Tirado et al. 2010; Chakauya et al. 2009) | Food supply instability is often driven by price volatility, which can be driven by rapid economic growth and which can contribute to consumer price inflation and higher import costs as a percentage of GDP leading to account deficits (Gilbert and Morgan 2010) | Excessive disruptions in food supply can place strains on infrastructure (e.g. needing additional storage facilities) (Yang and Zehnder 2002). Improved food transport can create demands for improved infrastructure (Akkerman et al. 2010; Shively and Thapa 2016). For example, weatherproofing transport systems and improving the efficiency of food trade (Ingram et al. 2016a; Stathers et al. 2013) | Food volatility makes it more challenging to supply food to vulnerable regions, and likely increases inequality (Baldos and Hertel 2015; Frank et al. 2017; Porter et al. 2014; Wheeler and von Braun 2013). Improved food distribution could reduce inequality in access to high quality nutritious foods. Food insecure consumers benefit from better access and distribution (e.g. food deserts) (Ingram 2011; | Improved food distribution can contribute to better food access and stronger urban communities (Kantor 2001; Hendrickson et al. 2006). Food price spikes often hit urban consumers the hardest in food importing countries, and increasing stability can reduce risk of food riots (Cohen and Garrett 2010) | Improved storage and distribution are likely to contribute to sustainable production by impacting biomass of paper/card and aluminum and iron-ore mining used for food packaging (Ingram et al. 2016a). | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | Better transport improves chances for expanding trade in developing countries (Newfarmer et al. 2009). Well-planned trade systems may act as a buffer to supply food to vulnerable regions (Baldos and Hertel 2015; Frank et al. 2017; Porter et al. 2014; Wheeler and von Braun 2013). |

| | | | | | | | | | | | | | | | | | | |
|--|---|--|---|---|---|--|--|---|---|--|--|---|--|--|--|--|---|--|
| | | 2008) to 450 million people (Brinkman et al. 2009), and caused welfare losses of 3% or more for poor households | food security in developing countries (Hine 1993). | | | | | | | especially in countries with inadequate infrastructure and weak food distribution systems (Vermeulen et al. 2012a), can strengthen climate resilience against future climate-related shocks (Ingram et al. 2016a; Stathers et al. 2013). | Coveney and O'Dwyer 2009) | | | | | | | |
| | | Regional food systems present opportunities for interconnectedness of the food system 's component resilient food supply systems and city-regions have an important role (Brinkley et al. 2016; Rocha 2016). However, mixed evidence on if urban agriculture contributes to poverty reduction (Ellis and Sumberg 1998) | Food insecurity in urban areas is often invisible (Crush and Frayne 2011). Improved urban food systems manage flows of food into, within, and out of the cities and have large role to play in reducing urban food security (Smit 2016; Benis and Ferrão 2017a; Brinkley et al. 2016; Rocha 2016; Maxwell and Wiebe 1999), particularly in fostering regional food self-reliance (Aldababseh et al. 2018; Bustamante et al. 2014b). | Since urban poor spend a great deal of their budget on food and urban diets are exposed to more unhealthy 'fast foods' (Dixon et al. 2007), local urban food systems can contribute to enhanced nutrition in urban areas (Tao et al. 2015; Maxwell 1999; Neff et al. 2009). However, local urban agriculture also may introduce pollution into food system through toxins in soil and water (Binns et al. 2003) | School feeding programs in urban areas can increase educational attendance and outcomes (Ashe and Sonnino 2013) | Urban and Peri-urban Agriculture and Forestry (UPAF) addresses gender-based differences in accessing food since women play an important role in the provisioning of urban food (Tao et al. 2015; Binns and Lynch 1998). Women also dominate informal urban food provisioning (wet markets, street food) (Smith 1998) | Water access often a constraint on urban agriculture (de Bon et al. 2010; Badami and Ramankutty 2015). Urban agriculture can exacerbate urban water pollution problems (pesticide runoff, etc) (Pothukuchi and Kaufman 1999) | Local food production and use can reduce energy use, due to lower demand of resources for production, transport and infrastructure (Lee-Smith 2010), but depends on context (Mariola 2008; Coley et al. 2009) | Urban food systems have as one aim to stimulate local economic development and increase employment in urban agriculture and food processing (Smith 1998). As many as 50% of some cities' retail jobs are in food-related sector (Pothukuchi and Kaufman 1999) | Urban food provisioning creates demands for expanded infrastructure in processing, refrigeration, and transportation (Pothukuchi and Kaufman 1999) | Many UFS in global South (e.g. Horizonte, Belo Horizonte, Brazil) have goals to reduce inequality in access to food. (Dixon et al. 2007; Allen 2010) | UFS aim at improving the health status of urban dwellers, reducing their exposure to pollution levels, and stimulating economic development (Tao et al. 2015) | UFS aim to combine sustainable production and consumption with local foodsheds (Tao et al. 2015; Allen 2010) | See main text on climate mitigation and adaptation | Overall reduced emissions would decrease rate of ocean acidification (Doney et al. 2009) | See main text on desertification and degradation | Building a resilient regional food system requires adjusting to the social and cultural environment and locally-specific natural resource base and building local institutions (Akhtar et al. 2016). Production of food within cities can potentially lead to less likelihood of urban food shortages and conflicts (Cohen & Garrett 2010). | N/A |
| | Improved food processing and retailing | Food processing has been a useful strategy for poverty reduction in some countries (Weinberger and Lumpkin 2007; Haggblade et al. 2010) | Efficiency in food processing and supply chains can contribute to more food reaching consumers and | Improved processing and distribution & storage systems can provide safer and healthier food to consumers (Vermeulen et | N/A | Improved food processing can displace street vendors and informal food sellers, who are predominantly women | Food processing and packaging activities such as washing, heating, cooling are heavily | Food processing activities such as heating and cooling are heavily dependent on energy so improved efficiency could reduce energy demand (Garcia and | Phytosanitary barriers currently prevent much food export from developing countries, and improvements in processing | Improvements in processing, refrigeration, and transportation will require investments in improved infrastructure | N/A | Improved food transport can reduce cities' ecological footprints and reduce overall emissions (Du et al. | Improved food processing and agro-retailing contributes to sustainable production (Ingram 2011) | See main text on climate mitigation and adaptation | Overall reduced emissions would decrease rate of ocean acidification (Doney et | See main text on desertification and degradation | N/A | Improved processing increases chances for expanding trade in developing countries (Newfarmer |

| | | | | | | | | | | | | | | | | | | |
|--|--|--|---|---|-----|--|---|--|---|---------------|-----|-------|---|--|---|--|-----|-----|
| | | | improved nutrition (Vermeulen et al. 2012a; Keding et al. 2013) | al. 2012a) and reduce food waste and health risks associated with poor storage management practices (James and James 2010a), although overpackaged prepared foods that are less healthy are also on rise (Monteiro 2009; Monteiro et al. 2011). | | (Smith 1998; Dixon et al. 2007) | dependent on freshwater so improved postharvest storage and distribution could reduce water demand via more efficiently performing systems (Garcia and You 2016). | You 2016). | would increase exports and GDP (Henson and Loader 2001; Jongwanich 2009). | (Ingram 2011) | | 2006) | | al. 2009) | | et al. 2009) | | |
| Improved energy use in food systems | Might possibly have impact on poverty by reducing farmer costs, but no data. | | Utilising energy-saving strategies can support reduced food waste (Ingram et al. 2016a) and increased production efficiencies (Smith and Gregory 2013). | Organic agriculture is associated with increased energy efficiency, which have can have co-benefits by reduced exposure to agrochemicals by farm workers (Gomiero et al. 2008) | N/A | Increased efficiency might reduce women's labor workloads on farms (Rahman 2010) but data is scarce. | Increased energy efficiency (e.g. in irrigation) can lead to more efficient water use (Rothausen and Conway 2011; Ringler and Lawford 2013) | Increased energy efficiency will reduce demands for energy but can have rebound effect in expanded acreage (Swanton et al. 1996) | There is no clear association between higher energy use in agriculture and economic growth; these have become decoupled in many countries (Bonny 1993). Data is unclear though on economic impacts of potential cost savings. | N/A | N/A | N/A | Reducing energy use in agriculture contributes to sustainable production goals (Ingram et al. 2016a). | See main text on climate mitigation and adaptation | Overall reduced emissions would decrease rate of ocean acidification (Doney et al. 2009). | See main text on desertification and degradation | N/A | N/A |

1
2
3
4

Table SM6.14 Impacts on the UN SDG of integrated response options based on risk management

| <u>Integrated response options based on risk management</u> | GOAL 1: No Poverty | GOAL 2: Zero Hunger | GOAL 3: Good Health and Well-being | GOAL 4: Quality Education | GOAL 5: Gender Equality | GOAL 6: Clean Water and Sanitation | GOAL 7: Affordable and Clean Energy | GOAL 8: Decent Work and Economic Growth | GOAL 9: Industry, Innovation and Infrastructure | GOAL 10: Reduced Inequality | GOAL 11: Sustainable Cities and Communities | GOAL 12: Responsible Consumption and Production | GOAL 13: Climate Action | GOAL 14: Life Below Water | GOAL 15: Life on Land | GOAL 16: Peace and Justice Strong Institutions | GOAL 17: Partnerships to achieve the Goal |
|---|---|--|--|---------------------------|-------------------------|--|---|---|--|---|--|---|--|---------------------------|--|--|---|
| Management of urban sprawl | Inner city poverty closely associated with urban sprawl in US context (Frumkin 2002; Powell 1999; Jargowsky | There are likely to be some benefits for food security since it is often agricultural land that is | Strong association between urban sprawl and poorer health outcomes | N/A | N/A | Urban sprawl is associated with higher levels of water pollution due to loss of filtering vegetation and increasing impervious | Sprawling or informal settlements often do not have access to electricity or other services, increasing | Sprawl is associated with rapid economic growth in some areas (Brueckner 2000). | Urban sprawl often increases public infrastructure costs (Brueckner 2000), and densification and | Urban sprawl is associated with inequality (Jargowsky 2002) | Urban sprawl is associated with unsustainability, including increased transport and CO ₂ emissions, lack of access to services, | Reducing urban sprawl and promoting community gardens and periurban agriculture can contribute to | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | There are debates over the role of urban sprawl in reducing social capital and weakening | N/A |

| | | | | | | | | | | | | | | | | | | |
|--|--|--|--|---|--|---|--|---|---|--|--|---|--|--|-----|--|---|--|
| | | 2002; Deng and Huang 2004) | sealed by the urban expansion (Barbero-Sierra et al. 2013a). Some evidence for sprawl reducing food production, particularly in China (Chen 2007b) | (air pollution, obesity, traffic accidents) (Frumkin 2002; Lopez 2004; Freudenberg et al. 2005) | | | surfaces (Romero and Ordenes 2004; Tu et al. 2007) | chances HH rely on dirty fuels (Dhingra et al. 2008) | Reducing urban sprawl is part of many managed "smart growth" plans, which may reduce overall economic growth in return for sustainability benefits (Godschalk 2003) | redevelopment can improve equality of access to infrastructure (Jenks and Burgess 2000). | | and loss of civic life (Kombe 2005; Andersson 2006). Sustainable cities include compactness, sustainable transport, density, mixed land uses, diversity, passive solar design, and greening (Chen et al. 2008; Jabareen 2006; Andersson 2006) | | | | participatory production in cities (Frumkin 2002; Nguyen 2010) | | |
| | | Diversification is associated with increased welfare and incomes and decreased levels of poverty in several country studies (Arslan et al. 2018b; Asfaw et al. 2018). | Diversification is associated with increased access to income and additional food sources for the household (Pretty 2003); likely some food security benefits but diversification can also lead to more purchased (unhealthy) foods (Niehof 2004; Barrett et al. 2001) | More diversified livelihoods have diversified diets which have better health outcomes (Block and Webb 2001; Kadiyala et al. 2014) particularly for women and children (Pretty 2003) | More diversified households tend to be more affluent, & have more income for education (Ellis 1998; Estudillo and Otsuka 1999; Steward 2007), but diversification through migration may reduce educational outcomes for children (Gioli et al. 2014) | Women are participants in and benefit from livelihood diversification, such as having increased control over sources of HH income (Smith 2015), although it can increase their labor requirements (Angeles and Hill 2009) | Lack of access to affordable water may inhibit livelihood diversification (Calow et al. 2010) | Access to clean energy can provide additional opportunities for livelihood diversification (Brew-Hammond 2010; Suckall et al. 2015) | Livelihood diversification by providing additional work opportunities (Ellis 1998; Niehof 2004) | N/A | The relationship between livelihood diversification and inequality is inconclusive (Ellis 1998). In some cases diversification on reduced inequality (Adams 1994) while it increases it (Reardon et al 2000) | One part of urban livelihoods in developing countries are linkages between rural and urban areas through migration and remittances (Rakodi 1999; Rakodi & Lloyd 2002); this livelihood diversification can strengthen urban income (Ricci 2012) | Livelihood diversification does not always lead to sustainable production and consumption choices, but it can strengthen autonomy potentially leading to better choices (Elmqvist and Olsson 2007; Schneider and Niederle 2010) | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | N/A | N/A |
| | | Many hundreds of millions of smallholders still rely on local seeds; without them they would have to find money to buy commercial seeds (Altieri et al. 2012b; McGuire and Sperling 2016; Howard 2015) | Local seeds revive and strengthen local food systems (McMichael and Schneider 2011b) and lead to more diverse and healthy food in areas with strong food sovereignty networks (Coomes et al. 2015a; Bisht et al. 2018). However local seeds often are less productive than improved varieties. | Local seed use is associated with fewer pesticides (Altieri et al. 2012b); loss of local seeds and substitution by commercial seeds is perceived by farmers to increase health risks (Mazzeo and Brenton 2013), although overall literature on links between food sovereignty and health is weak (Jones | N/A | Women play important roles in preserving and using local seeds (Ngoyea and Kumarakulasingam 2017; Bezner Kerr 2013) and sovereignty movements paying more attention to gender needs (Park et al. 2015) | Local seeds often have lower water demands, as well as less use of pesticides that can contaminate water (Adhikari 2014) | N/A | Food sovereignty supporters believe protecting smallholder agriculture provides more than commercial agriculture (Kloppenberg 2010) | N/A | Seed sovereignty advocates believe it will contribute to reduced inequality (Wittman 2011; Park et al. 2015) but there is inconclusive empirical evidence. | Seed sovereignty can help sustainable urban gardening (Demailly and Darly 2017) which can be part of a sustainable city by providing fresh, local food (Leitgeb et al. 2016). | Locally developed seeds can both help protect local agrobiodiversity and can often be more climate resilient than generic commercial varieties, leading to more sustainable production (Coomes et al. 2015a; van Niekerk and Wynberg 2017a). | See main text on climate mitigation and adaptation | N/A | See main text on desertification and degradation | Seed sovereignty is positively associated with strong local food movements, which contribute to social capital (McMichael and Schneider 2011b; Coomes et al. 2015a; Grey and Patel 2015). | Seed sovereignty could be seen as threat to free trade and imports of genetically modified seeds (Kloppenberg 2010; Howard 2015; Kloppenburg 2014) |

| | | | | | | | | | | | | | | | | | | |
|--|---------------------------------|---|--|---|--|---|--|-----|--|--|---|--|---|--|---|--|--|---|
| | | | | et al. 2015) | | | | | | | | | | | | | | |
| | Disaster risk management | DRM can help prevent impoverishment as disasters are a major factor in poverty (Basher 2006; Fothergill and Peck 2004) | Famine early warning systems have been successful to prevent impending food shortages (Genesio et al. 2011; Hillbruner and Moloney 2012) | EWS very important for public health to ensure people can get shelter and medical care during disasters (Greenough et al. 2001; Ebi and Schmier 2005) | N/A | Women often disproportionately affected by disasters; gender-sensitive EWS can reduce their vulnerability (Enarson and Meyreles 2004; Mustafa et al. 2015) | Many EWS include water monitoring components that contribute to access to clean water (Wilhite 2005; Iglesias et al. 2007). Some urban areas use water EWS successfully to monitor levels of contaminants (Hasan et al. 2009; Hou et al. 2013) | N/A | DRM can help minimise damage from disasters, which impacts economic growth (Basher 2006) | DRM can help protect infrastructures from damage during disaster (Rogers and Tsirkunov 2011) | EWS can ensure inequality is taken into account when making predictions of impacts (Khan et al. 1992) | EWS can be very effective in urban settings such as heat wave EWS and flooding EWS to minimise vulnerability (Parnell et al. 2007; Bambrick et al. 2011; Djordjević et al. 2011) | DRM can make sustainable production more possible by providing farmers with advance notice of environmental needs (Stigter et al. 2000; Parr et al. 2003) | See main text on climate mitigation and adaptation | EWS can play important role in marine management, e.g. warnings of red tide, tsunami warnings for coastal communities (Lee et al. 2005; Lauterjung et al. 2010) | See main text on desertification and degradation | DRM can reduce risk of conflict (Meier et al. 2007), increase resilience of communities (Mathbor 2007) and strengthen trust in institutions (Altieri et al. 2012b) | N/A |
| | Risk sharing instruments | Crop insurance reduces risks which can improve poverty outcomes by avoiding catastrophic losses, but is often not used by poorest people (Platteau et al. 2017) | Availability of crop insurance has generally lead to (modest) expansions in cultivated land area and increased food production (Claassen et al. 2011; Goodwin et al. 2004) | General forms of social protection lead to better health outcomes; unclear how much crop insurance contributes (Tirivayi et al. 2016) | Households lacking insurance may withdraw children from school after crop shocks (Jacoby and Skoufias 1997; Bandara et al. 2015) | Women farmers vulnerable to crop shocks, but tend to be more risk-averse and skeptical of commercial insurance (Aker et al. 2016; Fletschner and Kenney 2014) | Crop insurance can be indexed to weather and water access and thereby increase adaptation to water stress (Hoff and Bouwer 2003). Subsidised insurance can also be linked to reductions in pesticide use to reduce non-point source pollution, which has shown success in the US and China (Luo et al. 2014) | N/A | Subsidised crop insurance contributes to economic growth in the US (Atwood et al. 1996) but at considerable cost to the governance (Glauber 2004). | N/A | N/A | N/A | Crop insurance has been implicated as a driver of unsustainable production and disincentive to diversification (Bowman and Zilberman 2013), although community risk sharing might increase diversification and production | See main text on climate mitigation and adaptation | There is mixed evidence that crop insurance may encourage excess fertiliser use (Kramer et al. 1983; Wu 1999; Smith and Goodwin 1996), which contributes to ocean pollution; however, some governments are requiring reductions in nonpoint source pollution from farms otherwise farmers lose crop insurance (Iho et al. 2015) | See main text on desertification and degradation | Community risk sharing instruments can help strengthen resilience and institutions (Agrawal 2001) | Subsidised crop insurance can be seen as a subsidy and barrier to trade (Young and Westcott 2000) |

Supplementary Information for Section 6.4.4

| | IAM Study | C | M | A | D | L | F | O |
|--------------------------|-----------|-----|-----|-----|-----|-----|-----|-----|
| Alexander et al. 2018 | No | | | Yes | | | | Yes |
| Baker et al. 2019a | No | | Yes | | | | | |
| Baldos and Hertel 2014 | No | | | | | | Yes | |
| Bauer et al. 2018 | Yes | | Yes | | | | | |
| Bertram et al. 2018 | Yes | | Yes | | | | Yes | Yes |
| Brink et al. 2018 | Mixed | | | | Yes | Yes | Yes | Yes |
| Calvin et al. 2013b | Yes | | Yes | Yes | | | | |
| Calvin et al. 2014b | Yes | | Yes | | | | Yes | Yes |
| Calvin et al. 2016a | Yes | | Yes | | | | | |
| Calvin et al. 2016b | Yes | | Yes | | | | | |
| Calvin et al. 2017c | Yes | | Yes | | | | Yes | |
| Calvin et al. 2019 | Yes | | Yes | | | | | Yes |
| Chaturvedi et al. 2013 | Yes | | Yes | | | | | Yes |
| Clarke et al. 2014a | Yes | Yes | Yes | | | | | Yes |
| Collins et al. 2013 | No | Yes | | | | | | |
| Daioğlu et al. 2019 | Yes | | Yes | | | | | |
| Doelman et al. 2018 | Yes | | Yes | | | | Yes | |
| Edmonds et al. 2013b | Yes | | Yes | | | | | |
| Favero and Massetti 2014 | Yes | Yes | Yes | | | | | |
| Frank et al. 2015 | IAM-land | | Yes | | | | | |
| Frank et al. 2017 | Yes | | Yes | | | | Yes | |
| Fricko et al. 2017 | Yes | | Yes | | | | | |
| Fujimori et al. 2017b | Yes | | Yes | | | | | |
| Fujimori et al. 2018a | Yes | | Yes | | | | Yes | |
| Fujimori et al. 2019 | Mixed | | Yes | | | | Yes | |
| Gao and Bryan 2017b | No | | Yes | | | Yes | Yes | Yes |
| Graham et al. 2018b | Yes | | | | | | | Yes |
| Grubler et al. 2018 | Yes | | Yes | | | | Yes | Yes |
| Hanasaki et al. 2013b | Yes | | | | | | | Yes |
| Harrison et al. 2016 | Yes | | | | | | | Yes |
| Hasegawa et al. 2015a | Yes | | | | | | Yes | |
| Hasegawa et al. 2015b | Yes | | | | | | Yes | |
| Hasegawa et al. 2018 | Mixed | | | Yes | | | Yes | |
| Heck et al. 2018 | Mixed | Yes | Yes | | | | | Yes |
| Hejazi et al. 2014c | Yes | | Yes | | | | | Yes |
| Hejazi et al. 2015d | Yes | | Yes | | | | | Yes |
| Humpenöder et al. 2014 | Yes | | Yes | | | | | |
| Humpenöder et al. 2018b | IAM-land | | Yes | | | | Yes | Yes |
| Iyer et al. 2018 | Yes | | Yes | | | | Yes | Yes |
| Jones et al., 2013 | Yes | Yes | | | | | | |
| Jones et al. 2015 | Yes | | Yes | | | | | |
| Kim et al. 2016a | Yes | | | Yes | | | Yes | Yes |
| Kraxner et al. 2013 | No | | Yes | | | | | Yes |
| Kreidenweis et al. 2016a | Yes | | Yes | | | | Yes | |
| Kriegler et al. 2017 | Yes | | Yes | | | | Yes | |
| Kriegler et al. 2018a | Mixed | | Yes | | | | | |
| Kriegler et al. 2018b | Yes | | Yes | | | | | |
| Kyle et al. 2014 | Yes | | Yes | Yes | | | | |
| Lamontagne et al. 2018 | Yes | | Yes | | | | | |
| Le Page et al. 2013b | Yes | | Yes | | | | | |

| | | | | | | | | |
|--------------------------|----------|-----|-----|-----|-----|-----|-----|-----|
| Liu et al. 2017 | No | | | Yes | | | Yes | |
| Lotze-Campen et al. 2013 | Mixed | | | Yes | | | Yes | |
| Monier et al. 2018 | Yes | Yes | Yes | Yes | | | | Yes |
| Mouratiadou et al. 2016 | Yes | | Yes | | | | | Yes |
| Muratori et al. 2016 | Yes | | Yes | | | | Yes | |
| Nelson et al. 2014 | Mixed | | | Yes | | | Yes | |
| Newbold et al. 2015 | Mixed | | | | | | | Yes |
| Obersteiner et al. 2016b | IAM-land | | | | | | Yes | Yes |
| Parkinson et al. 2019 | Yes | | Yes | | | | | Yes |
| Patrizio et al. 2018 | No | | Yes | | | | | Yes |
| Pedercini et al. 2018 | No | | | | | | Yes | Yes |
| Pikaar et al. 2018 | IAM-land | | Yes | | | | | Yes |
| Popp et al. 2014a | Yes | | Yes | | | | | |
| Popp et al. 2017 | Yes | | Yes | | | | Yes | |
| Powers and Jetz 2019 | No | | | | | | | Yes |
| Riahi et al. 2017c | Yes | | Yes | | | | Yes | |
| Ringler et al. 2016 | Yes | | | Yes | | | Yes | Yes |
| Rogelj et al. 2018b | Yes | | Yes | | | | | |
| Springmann et al. 2018a | No | | Yes | | | | | Yes |
| Stehfest et al. 2019 | Mixed | | | | | | | |
| Stevanovic et al. 2016 | IAM-land | | | Yes | | | | |
| Stevanović et al. 2017 | IAM-land | | Yes | | | | Yes | |
| Tai et al. 2014 | No | | | | | | Yes | |
| Thornton et al. 2017 | Yes | Yes | Yes | Yes | | | Yes | |
| UNCCD 2017 | Mixed | | | | Yes | Yes | Yes | Yes |
| van Meijl et al. 2018 | Mixed | | Yes | Yes | | | Yes | |
| van Vuuren et al. 2015b | Yes | | Yes | | | | Yes | Yes |
| van Vuuren et al. 2017 | Yes | | Yes | | | | | |
| van Vuuren et al. 2018b | Yes | | Yes | | | | | |
| Weindl et al. 2015 | IAM-land | | | Yes | | | Yes | |
| Weindl et al. 2017 | IAM-land | | Yes | | | | | |
| Wiebe et al. 2015 | Mixed | | | Yes | | | Yes | |
| Wolff et al. 2018 | No | | | | Yes | Yes | | Yes |
| Wu et al. 2019 | Yes | | | | | | | |
| Yamagata et al. 2018 | No | | | | | Yes | | Yes |

References:

- Abass, A. B., G. Ndunguru, P. Mamiro, B. Alenkhe, N. Mlingi, and M. Bekunda, 2014: Post-harvest food losses in a maize-based farming system of semi-arid savannah area of Tanzania. *J. Stored Prod. Res.*, **57**, 49–57, doi:10.1016/J.JSPR.2013.12.004. <https://www.sciencedirect.com/science/article/pii/S0022474X1300101X> (Accessed April 15, 2019).
- Accorsi, R., A. Gallo, and R. Manzini, 2017: A climate driven decision-support model for the distribution of perishable products. *J. Clean. Prod.*, **165**, 917–929, doi:10.1016/j.jclepro.2017.07.170.
- Adamo, S. B., 2010: Environmental migration and cities in the context of global environmental change. *Curr. Opin. Environ. Sustain.*, **2**, 161–165, doi:10.1016/J.COSUST.2010.06.005. <https://www.sciencedirect.com/science/article/pii/S1877343510000503> (Accessed April 4, 2019).
- Adams, R. H., 1994: Non-farm income and inequality in rural Pakistan: A decomposition analysis. *J. Dev. Stud.*, **31**, 110–133, doi:10.1080/00220389408422350. <http://www.tandfonline.com/doi/abs/10.1080/00220389408422350> (Accessed April 15, 2019).
- Adger, W. N., 1999: Social vulnerability to climate change and extremes in coastal Vietnam. *World Dev.*, **27**, 249–269, doi:10.1016/S0305-750X(98)00136-3.
- Adger, W. N., 2009: Social Capital, Collective Action, and Adaptation to Climate Change. *Econ. Geogr.*, **79**, 387–404, doi:10.1111/j.1944-8287.2003.tb00220.x. <http://doi.wiley.com/10.1111/j.1944-8287.2003.tb00220.x> (Accessed November 7, 2018).
- Adger, W. N., N. Brooks, G. Bentham, and M. Agnew, 2004: *New indicators of vulnerability and adaptive capacity*. 128 pp.
- Adger, W. N., and Coauthors, 2011: Resilience implications of policy responses to climate change. *Wiley Interdiscip. Rev. Clim. Chang.*, **2**, 757–766, doi:10.1002/wcc.133. <http://doi.wiley.com/10.1002/wcc.133> (Accessed November 11, 2018).
- Adhikari, J., 2014: *Seed Sovereignty: Analysing the Debate on Hybrid Seeds and GMOs and Bringing About Sustainability in Agricultural Development*. https://www.foresaction.org/app/webroot/vendor/tinymce/editor/plugins/filemanager/files/JFL_VO12%281%29Adhikari.pdf (Accessed April 15, 2019).
- Adimassu, Z., A. Kessler, and L. Stroosnijder, 2013: Exploring co-investments in sustainable land management in the Central Rift Valley of Ethiopia. *Int. J. Sustain. Dev. World Ecol.*, **20**, 32–44, doi:10.1080/13504509.2012.740690. <https://www.tandfonline.com/doi/full/10.1080/13504509.2012.740690> (Accessed November 10, 2018).
- , S. Langan, and R. Johnston, 2016: Understanding determinants of farmers' investments in sustainable land management practices in Ethiopia: review and synthesis. *Environ. Dev. Sustain.*, **18**, 1005–1023, doi:10.1007/s10668-015-9683-5. <http://link.springer.com/10.1007/s10668-015-9683-5> (Accessed November 10, 2018).
- Affognon, H., C. Mutungi, P. Sanginga, and C. Borgemeister, 2015: Unpacking Postharvest Losses in Sub-Saharan Africa: A Meta-Analysis. *World Dev.*, **66**, 49–68, doi:10.1016/J.WORLDDEV.2014.08.002. <https://www.sciencedirect.com/science/article/pii/S0305750X14002307> (Accessed April 15, 2019).
- Aggarwal, P. K., and Coauthors, 2018: The climate-smart village approach: framework of an integrative strategy for scaling up adaptation options in agriculture. *Ecol. Soc.*, **23**, art14, doi:10.5751/ES-09844-230114. <https://www.ecologyandsociety.org/vol23/iss1/art14/> (Accessed June 2, 2018).
- Agrawal, A., 2001: Common property institutions and sustainable governance of resources. *World Dev.*,

- 29, 1649–1672. <https://www.sciencedirect.com/science/article/pii/S0305750X01000638> (Accessed November 10, 2018).
- Ahlgren, E. O., M. Börjesson Hagberg, and M. Grahn, 2017: Transport biofuels in global energy–economy modelling – a review of comprehensive energy systems assessment approaches. *GCB Bioenergy*, doi:10.1111/gcbb.12431.
- Ahmed, S., and J. R. Stepp, 2016a: Beyond yields: Climate change effects on specialty crop quality and agroecological management. *Elem. Sci. Anthr.*, **4**, 92, doi:10.12952/journal.elementa.000092. <https://www.elementascience.org/articles/10.12952/journal.elementa.000092> (Accessed May 31, 2018).
- , and ———, 2016b: Beyond yields: Climate change effects on specialty crop quality and agroecological management. *Elem. Sci. Anthr.*, **4**, 92, doi:10.12952/journal.elementa.000092. <https://www.elementascience.org/articles/10.12952/journal.elementa.000092>.
- Ajibade, I., and G. McBean, 2014: Climate extremes and housing rights: A political ecology of impacts, early warning and adaptation constraints in Lagos slum communities. *Geoforum*, **55**, 76–86, doi:10.1016/j.geoforum.2014.05.005.
- Akhtar, P., Y. Tse, Z. Khan, and R. Rao-Nicholson, 2016: Data-driven and adaptive leadership contributing to sustainability: Global agri-food supply chains connected with emerging markets. *Int. J. Prod. Econ.*, **181**, 392–401. <https://www.sciencedirect.com/science/article/pii/S0925527315005125> (Accessed May 1, 2018).
- Akkerman, R., P. Farahani, and M. Grunow, 2010: Quality, safety and sustainability in food distribution: a review of quantitative operations management approaches and challenges. *OR Spectr.*, **32**, 863–904, doi:10.1007/s00291-010-0223-2. <http://link.springer.com/10.1007/s00291-010-0223-2> (Accessed April 15, 2019).
- Akter, S., T. J. Krupnik, F. Rossi, and F. Khanam, 2016: The influence of gender and product design on farmers’ preferences for weather-indexed crop insurance. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2016.03.010.
- Al-Mansour F, and Jecic V, 2017: A model calculation of the carbon footprint of agricultural products: The case of Slovenia. *Energy*, **136**, 7–15. <https://www.sciencedirect.com/science/article/pii/S0360544216315444> (Accessed November 11, 2018).
- Albert, C., T. Zimmermann, J. Knieling, and C. von Haaren, 2012: Social learning can benefit decision-making in landscape planning: Gartow case study on climate change adaptation, Elbe valley biosphere reserve. *Landsc. Urban Plan.*, **105**, 347–360, doi:10.1016/j.landurbplan.2011.12.024.
- Albrecht, A., and S. T. Kandji, 2003: Carbon sequestration in tropical agroforestry systems. *Agric. Ecosyst. Environ.*, **99**, 15–27, doi:10.1016/S0167-8809(03)00138-5. <https://www.sciencedirect.com/science/article/pii/S0167880903001385> (Accessed April 12, 2019).
- Aldababseh, A., M. Temimi, and P. Maghelal, 2018: Multi-Criteria Evaluation of Irrigated Agriculture Suitability to Achieve Food Security in an Arid Environment. *Sustainability*, **10**, 803–836, doi:10.3390/su10030803.
- Alderman, H., 2010: Safety Nets Can Help Address the Risks to Nutrition from Increasing Climate Variability. *J. Nutr.*, **140**, 148S–152S, doi:10.3945/jn.109.110825. <https://academic.oup.com/jn/article/140/1/148S/4600294> (Accessed November 11, 2018).
- Aleksandrowicz, L., R. Green, E. J. M. Joy, P. Smith, and A. Haines, 2016: The Impacts of Dietary Change on Greenhouse Gas Emissions, Land Use, Water Use, and Health: A Systematic Review. *PLoS One*, **11**, e0165797, doi:10.1371/journal.pone.0165797.

- <https://dx.plos.org/10.1371/journal.pone.0165797> (Accessed November 11, 2018).
- Alessa, L., A. Kliskey, J. Gamble, M. Fidel, G. Beaujean, and J. Gosz, 2016: The role of Indigenous science and local knowledge in integrated observing systems: moving toward adaptive capacity indices and early warning systems. *Sustain. Sci.*, **11**, 91–102, doi:10.1007/s11625-015-0295-7. <http://link.springer.com/10.1007/s11625-015-0295-7> (Accessed December 26, 2017).
- Alexander, P., M. D. A. Rounsevell, C. Dislich, J. R. Dodson, K. Engström, and D. Moran, 2015: Drivers for global agricultural land use change: The nexus of diet, population, yield and bioenergy. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2015.08.011.
- , C. Brown, A. Arneth, J. Finnigan, and M. D. A. Rounsevell, 2016: Human appropriation of land for food: The role of diet. *Glob. Environ. Chang.*, **41**, 88–98, doi:10.1016/J.GLOENVCHA.2016.09.005. <https://www.sciencedirect.com/science/article/pii/S0959378016302370> (Accessed April 5, 2019).
- , S. Rabin, P. Anthoni, R. Henry, T. A. M. Pugh, M. D. A. Rounsevell, and A. Arneth, 2018: Adaptation of global land use and management intensity to changes in climate and atmospheric carbon dioxide. *Glob. Chang. Biol.*, **24**, 2791–2809, doi:10.1111/gcb.14110. <http://doi.wiley.com/10.1111/gcb.14110> (Accessed April 14, 2019).
- Ali, G., N. Pumijumnong, and S. Cui, 2018: Valuation and validation of carbon sources and sinks through land cover/use change analysis: The case of Bangkok metropolitan area. *Land use policy*, **70**, 471–478, doi:10.1016/J.LANDUSEPOL.2017.11.003. <https://www.sciencedirect.com/science/article/pii/S0264837717311316> (Accessed May 31, 2018).
- Alkama, R., and A. Cescatti, 2016: Biophysical climate impacts of recent changes in global forest cover. *Science (80-.)*, **351**, 600–604, doi:10.1126/science.aac8083.
- Allen, P., 2010: Realizing justice in local food systems. *Cambridge J. Reg. Econ. Soc.*, **3**, 295–308, doi:10.1093/cjres/rsq015. <https://academic.oup.com/cjres/article-lookup/doi/10.1093/cjres/rsq015> (Accessed April 15, 2019).
- Allington, G., T. V.-J. of A. Environments, and undefined 2010, Reversal of desertification: the role of physical and chemical soil properties. *Elsevier*.
- Altieri, M. A., and C. I. Nicholls, 2017: The adaptation and mitigation potential of traditional agriculture in a changing climate. *Clim. Change*, **140**, 33–45, doi:10.1007/s10584-013-0909-y.
- , F. R. Funes-Monzote, and P. Petersen, 2012a: Agroecologically efficient agricultural systems for smallholder farmers: contributions to food sovereignty. *Agron. Sustain. Dev.*, **32**, 1–13, doi:10.1007/s13593-011-0065-6. <http://link.springer.com/10.1007/s13593-011-0065-6> (Accessed April 15, 2019).
- , ———, and ———, 2012b: Agroecologically efficient agricultural systems for smallholder farmers: contributions to food sovereignty. *Agron. Sustain. Dev.*, **32**, 1–13, doi:10.1007/s13593-011-0065-6. <http://link.springer.com/10.1007/s13593-011-0065-6> (Accessed November 11, 2018).
- Altieri, M. A., C. I. Nicholls, A. Henao, and M. A. Lana, 2015: Agroecology and the design of climate change-resilient farming systems. *Agron. Sustain. Dev.*, **35**, 869–890, doi:10.1007/s13593-015-0285-2. <http://link.springer.com/10.1007/s13593-015-0285-2> (Accessed June 1, 2018).
- Altman, M., T. G. Hart, and P. T. Jacobs, 2009: Household food security status in South Africa. *Agrekon*, **48**, 345–361, doi:10.1080/03031853.2009.9523831. <http://www.tandfonline.com/doi/abs/10.1080/03031853.2009.9523831> (Accessed April 15, 2019).
- Ampaire, E. L., L. Jassogne, H. Providence, M. Acosta, J. Twyman, L. Winowiecki, and P. van Asten, 2017: Institutional challenges to climate change adaptation: A case study on policy action gaps in

- Uganda. *Environ. Sci. Policy*, **75**, 81–90, doi:10.1016/j.envsci.2017.05.013.
- Anderson, C. M., C. B. Field, and K. J. Mach, 2017: Forest offsets partner climate-change mitigation with conservation. *Front. Ecol. Environ.*, **15**, 359–365, doi:10.1002/fee.1515. <http://doi.wiley.com/10.1002/fee.1515> (Accessed May 1, 2018).
- Anderson, H. R., 2017: Implications for the science of air pollution and health. *Lancet Respir. Med.*, **5**, 916–918, doi:10.1016/S2213-2600(17)30396-X.
- Anderson, K., and G. Peters, 2016: The trouble with negative emissions. *Science (80-.)*, **354**, 182–183, doi:10.1126/science.aah4567.
- Andersson, E., 2006: Urban Landscapes and Sustainable Cities. *Ecol. Soc.*, **11**, art34, doi:10.5751/ES-01639-110134. <http://www.ecologyandsociety.org/vol11/iss1/art34/> (Accessed April 15, 2019).
- Anenberg, S., and Coauthors, 2012: Global air quality and health co-benefits of mitigating near-term climate change through methane and black carbon emission controls. *Environ. Health Perspect.*, **120**, 831–839.
- Angeles, L. C., and K. Hill, 2009: The gender dimension of the agrarian transition: women, men and livelihood diversification in two peri-urban farming communities in the Philippines. *Gender, Place Cult.*, **16**, 609–629, doi:10.1080/09663690903148465. <http://www.tandfonline.com/doi/full/10.1080/09663690903148465> (Accessed April 15, 2019).
- Angelsen, A., 2010: Climate Mitigation and Food Production in Tropical Landscapes Special Feature: Policies for reduced deforestation and their impact on agricultural production. *Proc. Natl. Acad. Sci. U. S. A.*, **107**, 19639–19644.
- Annan, F., and W. Schlenker, 2015: Federal Crop Insurance and the Disincentive to Adapt to Extreme Heat. *Am. Econ. Rev.*, **105**, 262–266, doi:10.1257/aer.p20151031. <http://pubs.aeaweb.org/doi/10.1257/aer.p20151031> (Accessed June 1, 2018).
- Ansah, I. G. K., B. K. D. Tetteh, and S. A. Donkoh, 2017: Determinants and income effect of yam postharvest loss management: evidence from the Zabzugu District of Northern Ghana. *Food Secur.*, **9**, 611–620, doi:10.1007/s12571-017-0675-1. <http://link.springer.com/10.1007/s12571-017-0675-1> (Accessed May 10, 2018).
- Antille, D. L., J. M. Bennett, and T. A. Jensen, 2016a: Soil compaction and controlled traffic considerations in Australian cotton-farming systems. *Crop Pasture Sci.*, **67**, 1–28, doi:10.1071/CP15097. <http://www.publish.csiro.au/?paper=CP15097> (Accessed May 31, 2018).
- , ———, and ———, 2016b: Soil compaction and controlled traffic considerations in Australian cotton-farming systems. *Crop Pasture Sci.*, **67**, 1, doi:10.1071/CP15097. <http://www.publish.csiro.au/?paper=CP15097> (Accessed April 18, 2019).
- Antwi-Agyei, P., L. C. Stringer, and A. J. Dougill, 2014: Livelihood adaptations to climate variability: insights from farming households in Ghana. *Reg. Environ. Chang.*, **14**, 1615–1626, doi:10.1007/s10113-014-0597-9. <http://link.springer.com/10.1007/s10113-014-0597-9> (Accessed June 1, 2018).
- Archer, S., K. Davies, T. Fulbright, K. McDaniel, B. Wilcox, K. Predick, and D. Briske, 2011: Brush management as a rangeland conservation strategy: a critical evaluation. *Conservation benefits of rangeland practices: assessment, recommendations, and knowledge gaps.*, USDA-NRCS, Washington DC, 105–170 <https://www.ars.usda.gov/research/publications/publication/?seqNo115=268913> (Accessed November 11, 2018).
- Armitage, D., A. Dzyundzyak, J. Baird, Ö. Bodin, R. Plummer, and L. Schultz, 2018: An Approach to

- Assess Learning Conditions, Effects and Outcomes in Environmental Governance. *Environ. Policy Gov.*, **28**, 3–14, doi:10.1002/eet.1781. <http://doi.wiley.com/10.1002/eet.1781> (Accessed November 10, 2018).
- Arnáez J, Lana-Renault N, Lasanta T, Ruiz-Flaño P, and Castroviejo J, 2015: Effects of farming terraces on hydrological and geomorphological processes. A review. *Catena*, **128**, 122–134. <https://www.sciencedirect.com/science/article/pii/S0341816215000351> (Accessed November 11, 2018).
- Arndt, C., M. Hussain, and L. Østerdal, 2012: *Effects of food price shocks on child malnutrition: The Mozambican experience 2008/09*. <https://www.econstor.eu/handle/10419/80907> (Accessed November 9, 2018).
- Arndt, C., M. A. Hussain, V. Salvucci, and L. P. Østerdal, 2016: Effects of food price shocks on child malnutrition: The Mozambican experience 2008/2009. *Econ. Hum. Biol.*, **22**, 1–13, doi:10.1016/J.EHB.2016.03.003. <https://www.sciencedirect.com/science/article/pii/S1570677X16300119> (Accessed June 1, 2018).
- Arneeth, A., and Coauthors, 2010: From biota to chemistry and climate: towards a comprehensive description of trace gas exchange between the biosphere and atmosphere. *Biogeosciences*, **7**, 121–149, doi:10.5194/bg-7-121-2010. <http://www.biogeosciences.net/7/121/2010/> (Accessed June 2, 2018).
- Arora, V. K., and A. Montenegro, 2011: Small temperature benefits provided by realistic afforestation efforts. *Nat. Geosci.*, **4**, 514–518, doi:10.1038/ngeo1182.
- , and J. R. Melton, 2018: Reduction in global area burned and wildfire emissions since 1930s enhances carbon uptake by land. *Nat. Commun.*, **9**, 1326, doi:10.1038/s41467-018-03838-0. <http://www.nature.com/articles/s41467-018-03838-0> (Accessed November 10, 2018).
- Arslan, A., R. Cavatassi, F. Alfani, N. McCarthy, L. Lipper, and M. Kokwe, 2018a: Diversification Under Climate Variability as Part of a CSA Strategy in Rural Zambia. *J. Dev. Stud.*, **54**, 457–480, doi:10.1080/00220388.2017.1293813. <https://www.tandfonline.com/doi/full/10.1080/00220388.2017.1293813> (Accessed April 15, 2019).
- , ———, ———, ———, ———, and ———, 2018b: Diversification Under Climate Variability as Part of a CSA Strategy in Rural Zambia. *J. Dev. Stud.*, **54**, 457–480, doi:10.1080/00220388.2017.1293813. <https://www.tandfonline.com/doi/full/10.1080/00220388.2017.1293813> (Accessed June 1, 2018).
- Asfaw, S., G. Pallante, and A. Palma, 2018: Diversification Strategies and Adaptation Deficit: Evidence from Rural Communities in Niger. *World Dev.*, **101**, 219–234, doi:10.1016/J.WORLDDEV.2017.09.004. <https://www.sciencedirect.com/science/article/pii/S0305750X16301863> (Accessed June 1, 2018).
- Ashe, L. M., and R. Sonnino, 2013: At the crossroads: new paradigms of food security, public health nutrition and school food. *Public Health Nutr.*, **16**, 1020–1027, doi:10.1017/S1368980012004326. http://www.journals.cambridge.org/abstract_S1368980012004326 (Accessed April 15, 2019).
- Atwood, J. A., M. J. Watts, and A. E. Baquet, 1996: An Examination of the Effects of Price Supports and Federal Crop Insurance Upon the Economic Growth, Capital Structure, and Financial Survival of Wheat Growers in the Northern High Plains. *Am. J. Agric. Econ.*, **78**, 212–224, doi:10.2307/1243792. <https://academic.oup.com/ajae/article-lookup/doi/10.2307/1243792> (Accessed April 15, 2019).
- Avetisyan, M., T. Hertel, and G. Sampson, 2014: Is Local Food More Environmentally Friendly? The GHG Emissions Impacts of Consuming Imported versus Domestically Produced Food. *Environ. Resour. Econ.*, **58**, 415–462, doi:10.1007/s10640-013-9706-3.

- Ayers, J., and T. Forsyth, 2009: Community-Based Adaptation to Climate Change. *Environ. Sci. Policy Sustain. Dev.*, doi:10.3200/ENV.51.4.22-31.
- , and D. Dodman, 2010: Climate change adaptation and development I: The state of the debate. *Prog. Dev. Stud.*, **10**, 161–168, doi:10.1177/146499340901000205.
- Baccini, A., W. Walker, L. Carvalho, M. Farina, D. Sulla-Menashe, and R. A. Houghton, 2017: Tropical forests are a net carbon source based on aboveground measurements of gain and loss. *Science*, **358**, 230–234, doi:10.1126/science.aam5962. <http://www.ncbi.nlm.nih.gov/pubmed/28971966> (Accessed April 5, 2019).
- Badami, M. G., and N. Ramankutty, 2015: Urban agriculture and food security: A critique based on an assessment of urban land constraints. *Glob. Food Sec.*, **4**, 8–15, doi:10.1016/J.GFS.2014.10.003. <https://www.sciencedirect.com/science/article/abs/pii/S2211912414000431> (Accessed April 15, 2019).
- Bailey, R., 2013: The “Food Versus Fuel” Nexus. *The Handbook of Global Energy Policy*.
- Bailis, R., R. Drigo, A. Ghilardi, O. M.-N. C. Change, and undefined 2015, The carbon footprint of traditional woodfuels. *nature.com*,
- Bajželj, B., K. S. Richards, J. M. Allwood, P. Smith, J. S. Dennis, E. Curmi, and C. A. Gilligan, 2014: Importance of food-demand management for climate mitigation. *Nat. Clim. Chang.*, **4**, 924–929, doi:10.1038/nclimate2353. <https://www.nature.com/articles/doi:10.1038%2Fncclimate2353> (Accessed May 1, 2018).
- Baker, J. S., C. M. Wade, B. L. Sohngen, S. Ohrel, and A. A. Fawcett, 2019a: Potential complementarity between forest carbon sequestration incentives and biomass energy expansion. *Energy Policy*, **126**, 391–401, doi:10.1016/J.ENPOL.2018.10.009. <https://www.sciencedirect.com/science/article/pii/S030142151830661X> (Accessed April 5, 2019).
- Baker, J. S., C. M. Wade, B. L. Sohngen, S. Ohrel, and A. A. Fawcett, 2019b: Potential complementarity between forest carbon sequestration incentives and biomass energy expansion. *Energy Policy*, **126**, 391–401, doi:10.1016/j.enpol.2018.10.009. <https://doi.org/10.1016/j.enpol.2018.10.009>.
- Bala, G., N. Devaraju, R. K. Chaturvedi, K. Caldeira, and R. Nemani, 2013: Nitrogen deposition: how important is it for global terrestrial carbon uptake? *Biogeosciences*, **10**, 7147–7160, doi:10.5194/bg-10-7147-2013. <https://www.biogeosciences.net/10/7147/2013/> (Accessed November 12, 2018).
- Baldos, U. L. C., and T. W. Hertel, 2014a: Global food security in 2050: The role of agricultural productivity and climate change. *Aust. J. Agric. Resour. Econ.*, **58**, 554–570, doi:10.1111/1467-8489.12048.
- Baldos, U. L. C., and T. W. Hertel, 2014b: Global food security in 2050: the role of agricultural productivity and climate change. *Aust. J. Agric. Resour. Econ.*, **58**, 554–570, doi:10.1111/1467-8489.12048. <http://doi.wiley.com/10.1111/1467-8489.12048> (Accessed April 5, 2019).
- , and ———, 2015: The role of international trade in managing food security risks from climate change. *Food Secur.*, **7**, 275–290, doi:10.1007/s12571-015-0435-z. <http://link.springer.com/10.1007/s12571-015-0435-z> (Accessed April 4, 2019).
- Balmford, A., and Coauthors, 2018: The environmental costs and benefits of high-yield farming. *Nat. Sustain.*, **1**, 477. <https://www.nature.com/articles/s41893-018-0138-5> (Accessed November 10, 2018).
- Bambrick, H. J., A. G. Capon, G. B. Barnett, R. M. Beaty, and A. J. Burton, 2011: Climate Change and Health in the Urban Environment: Adaptation Opportunities in Australian Cities. *Asia Pacific J. Public Heal.*, **23**, 67S–79S, doi:10.1177/1010539510391774.

- <http://journals.sagepub.com/doi/10.1177/1010539510391774> (Accessed April 15, 2019).
- Bamman, H., 2007: *Participatory value chain analysis for improved farmer incomes, employment opportunities and food security*. 1-13 pp.
- Bandara, A., R. Dehejia, and S. Lavie-Rouse, 2015: The Impact of Income and Non-Income Shocks on Child Labor: Evidence from a Panel Survey of Tanzania. *World Dev.*, **67**, 218–237, doi:10.1016/J.WORLDDEV.2014.10.019. <https://www.sciencedirect.com/science/article/abs/pii/S0305750X14003301> (Accessed April 15, 2019).
- Baptista, F., L. L. Silva, C. De Visser, J. Gołaszewski, A. Meyer-Aurich, D. Briassoulis, H. Mikkola, and D. Murcho, 2013: Energy Efficiency in Agriculture. *Complete communications of the 5th International Congress on Energy and Environment Engineering and Management* https://dspace.uevora.pt/rdpc/bitstream/10174/8648/1/Energy_efficiency_in_agriculture_Lisbon.pdf (Accessed June 1, 2018).
- Barbero-Sierra, C., M. J. Marques, and M. Ruiz-Pérez, 2013a: The case of urban sprawl in Spain as an active and irreversible driving force for desertification. *J. Arid Environ.*, **90**, 95–102, doi:10.1016/j.jaridenv.2012.10.014. <https://linkinghub.elsevier.com/retrieve/pii/S0140196312002820> (Accessed November 11, 2018).
- , ———, and ———, 2013b: The case of urban sprawl in Spain as an active and irreversible driving force for desertification. *J. Arid Environ.*, **90**, 95–102, doi:10.1016/j.jaridenv.2012.10.014. <https://linkinghub.elsevier.com/retrieve/pii/S0140196312002820> (Accessed November 10, 2018).
- Bárcena, T. G., L. P. Kiær, L. Vesterdal, H. M. Stefánsdóttir, P. Gundersen, and B. D. Sigurdsson, 2014: Soil carbon stock change following afforestation in Northern Europe: A meta-analysis. *Glob. Chang. Biol.*, doi:10.1111/gcb.12576.
- Barlow, J., and Coauthors, 2016: Anthropogenic disturbance in tropical forests can double biodiversity loss from deforestation. *Nature*, **535**, 144–147, doi:10.1038/nature18326. <http://www.nature.com/articles/nature18326> (Accessed November 11, 2018).
- Barnes, A. P., and S. G. Thomson, 2014a: Measuring progress towards sustainable intensification: How far can secondary data go? *Ecol. Indic.*, **36**, 213–220, doi:10.1016/j.ecolind.2013.07.001.
- Barnes, A. P., and S. G. Thomson, 2014b: Measuring progress towards sustainable intensification: How far can secondary data go? *Ecol. Indic.*, **36**, 213–220, doi:10.1016/j.ecolind.2013.07.001. <https://linkinghub.elsevier.com/retrieve/pii/S1470160X1300263X> (Accessed April 18, 2019).
- Barnes, A. P., H. Hansson, G. Manevska-Tasevska, S. Shrestha, and S. G. Thomson, 2015: The influence of diversification on long-term viability of the agriculture sector. *Land use policy*, **49**, 404–412, doi:10.1016/j.landusepol.2015.08.023. <http://creativecommons.org/licenses/by-nc-nd/4.0/> (Accessed May 31, 2018).
- Barnett, J., and J. Palutikof, 2015: 26 The limits to adaptation. A comparative analysis. *Applied Studies in Climate Adaptation*, John Wiley & Sons, Ltd., p. 231.
- Barrett, C., T. Reardon, and P. Webb, 2001: Nonfarm income diversification and household livelihood strategies in rural Africa: concepts, dynamics, and policy implications. *Food Policy*, **26**, 315–331. <https://www.sciencedirect.com/science/article/pii/S0306919201000148> (Accessed November 11, 2018).
- Barthel, S., and C. Isendahl, 2013: Urban gardens, agriculture, and water management: Sources of resilience for long-term food security in cities. *Ecol. Econ.*, **86**, 224–234, doi:10.1016/J.ECOLECON.2012.06.018. <https://www.sciencedirect.com/science/article/pii/S0921800912002431> (Accessed June 1, 2018).

- Basher, R., 2006: Global early warning systems for natural hazards: systematic and people-centred. *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, doi:10.1098/rsta.2006.1819.
- Bassoum, S., and D. Ghiggi, 2010: Sahel vert: a project of Centre Mampuya, Senegal. *In International Symposium on Urban and Peri-Urban Horticulture in the Century of Cities: Lessons, Challenges, Opportunitites 1021*, 367–372 https://www.actahort.org/books/1021/1021_33.htm (Accessed November 11, 2018).
- Bastin, J.-F., and Coauthors, 2017: The extent of forest in dryland biomes. *Science*, **356**, 635–638, doi:10.1126/science.aam6527. <http://www.ncbi.nlm.nih.gov/pubmed/28495750> (Accessed April 5, 2019).
- Batidzirai, B., E. M. W. Smeets, and A. P. C. Faaij, 2012: Harmonising bioenergy resource potentials—Methodological lessons from review of state of the art bioenergy potential assessments. *Renew. Sustain. Energy Rev.*, **16**, 6598–6630, doi:https://doi.org/10.1016/j.rser.2012.09.002. <http://www.sciencedirect.com/science/article/pii/S1364032112004996>.
- Batterbury, S., 2001: Landscapes of Diversity: A Local Political Ecology of Livelihood Diversification in South-Western Niger. *Ecumene*, **8**, 437–464, doi:10.1177/096746080100800404. <http://journals.sagepub.com/doi/10.1177/096746080100800404> (Accessed November 11, 2018).
- Baudron, F., and Coauthors, 2015a: Re-examining appropriate mechanization in Eastern and Southern Africa: two-wheel tractors, conservation agriculture, and private sector involvement. *Food Secur.*, **7**, 889–904, doi:10.1007/s12571-015-0476-3. <http://link.springer.com/10.1007/s12571-015-0476-3> (Accessed June 1, 2018).
- , and Coauthors, 2015b: Re-examining appropriate mechanization in Eastern and Southern Africa: two-wheel tractors, conservation agriculture, and private sector involvement. *Food Secur.*, **7**, 889–904, doi:10.1007/s12571-015-0476-3. <http://link.springer.com/10.1007/s12571-015-0476-3> (Accessed April 18, 2019).
- Bauer, N., and Coauthors, 2018: Global energy sector emission reductions and bioenergy use: overview of the bioenergy demand phase of the EMF-33 model comparison. *Clim. Change*, 1–16, doi:10.1007/s10584-018-2226-y. <http://link.springer.com/10.1007/s10584-018-2226-y> (Accessed April 14, 2019).
- Baumhardt, R. L., B. A. Stewart, and U. M. Sainju, 2015: North American soil degradation: Processes, practices, and mitigating strategies. *Sustain.*, **7**, 2936–2960, doi:10.3390/su7032936.
- Baur, A. H., M. Förster, and B. Kleinschmit, 2015: The spatial dimension of urban greenhouse gas emissions: analyzing the influence of spatial structures and LULC patterns in European cities. *Landsc. Ecol.*, **30**, 1195–1205, doi:10.1007/s10980-015-0169-5. <http://link.springer.com/10.1007/s10980-015-0169-5> (Accessed August 11, 2018).
- Bautista, S., J. Llovet, A. Ocampo-Melgar, A. Vilagrosa, Á. Mayor, C. Murias, V. Vallejo, and B. Orr, 2017: Integrating knowledge exchange and the assessment of dryland management alternatives—A learning-centered participatory approach. *J. Environ. Manage.*, **195**, 35–45. <https://www.sciencedirect.com/science/article/pii/S0301479716309380> (Accessed November 10, 2018).
- Baveye, P. C., J. Berthelin, D. Tessier, and G. Lemaire, 2018a: The “4 per 1000” initiative: A credibility issue for the soil science community? *Geoderma*, **309**, 118–123, doi:10.1016/j.geoderma.2017.05.005. <http://linkinghub.elsevier.com/retrieve/pii/S0016706117306171> (Accessed May 30, 2018).
- , ———, ———, and ———, 2018b: The “4 per 1000” initiative: A credibility issue for the soil science community? *Geoderma*, **309**, 118–123, doi:10.1016/j.geoderma.2017.05.005.

- <https://linkinghub.elsevier.com/retrieve/pii/S0016706117306171> (Accessed April 18, 2019).
- Bayrak, M., L. Marafa, M. M. Bayrak, and L. M. Marafa, 2016: Ten Years of REDD+: A Critical Review of the Impact of REDD+ on Forest-Dependent Communities. *Sustainability*, **8**, 620, doi:10.3390/su8070620. <http://www.mdpi.com/2071-1050/8/7/620> (Accessed April 5, 2019).
- Beerling, D. J., and Coauthors, 2018: Farming with crops and rocks to address global climate, food and soil security. *nature.com*.
- Begum, R., K. Sohag, S. Abdullah, and M. Jaafar, 2015: CO2 emissions, energy consumption, economic and population growth in Malaysia. *Renew. Sustain. Energy Rev.*, **41**, 594–601. <https://www.sciencedirect.com/science/article/pii/S1364032114006650> (Accessed November 11, 2018).
- Begum, R. A., M. S. K. Sarkar, A. H. Jaafar, and J. J. Pereira, 2014: Toward conceptual frameworks for linking disaster risk reduction and climate change adaptation. *Int. J. Disaster Risk Reduct.*, **10**, 362–373, doi:10.1016/J.IJDRR.2014.10.011. <https://www.sciencedirect.com/science/article/pii/S2212420914000971> (Accessed April 4, 2019).
- Behrman, K. D., T. E. Juenger, J. R. Kiniry, and T. H. Keitt, 2015: Spatial land use trade-offs for maintenance of biodiversity, biofuel, and agriculture. *Landsc. Ecol.*, doi:10.1007/s10980-015-0225-1.
- Bello, C., and Coauthors, 2015: Defaunation affects carbon storage in tropical forests. *Sci. Adv.*, **1**, e1501105, doi:10.1126/sciadv.1501105. <http://advances.sciencemag.org/lookup/doi/10.1126/sciadv.1501105> (Accessed April 19, 2019).
- Benis, K., and P. Ferrão, 2017a: Potential mitigation of the environmental impacts of food systems through urban and peri-urban agriculture (UPA)—a life cycle assessment approach. *J. Clean. Prod.*, **140**, 784–795. <https://www.sciencedirect.com/science/article/pii/S0959652616306552> (Accessed May 1, 2018).
- Bejamin, E.O., Ola, O. & Buchenrieder, G. 2018. Does an agroforestry scheme with payment for ecosystem services (PES) economically empower women in sub-Saharan Africa? *Ecosystem Services* 31, 1-11
- , and P. Ferrão, 2017b: Potential mitigation of the environmental impacts of food systems through urban and peri-urban agriculture (UPA)—a life cycle assessment approach. *J. Clean. Prod.*, **140**, 784–795. <https://www.sciencedirect.com/science/article/pii/S0959652616306552> (Accessed May 1, 2018).
- Bennetzen, E., P. Smith, J. P.-G. E. Change, and undefined 2016, Agricultural production and greenhouse gas emissions from world regions—the major trends over 40 years. *Elsevier*.
- Bennetzen, E. H., P. Smith, and J. R. Porter, 2016: Decoupling of greenhouse gas emissions from global agricultural production: 1970-2050. *Glob. Chang. Biol.*, **22**, 763–781, doi:10.1111/gcb.13120. <http://doi.wiley.com/10.1111/gcb.13120> (Accessed November 8, 2018).
- Berman, R., C. Quinn, and J. Paavola, 2012a: The role of institutions in the transformation of coping capacity to sustainable adaptive capacity. *Environ. Dev.*, **2**, 86–100, doi:10.1016/j.envdev.2012.03.017.
- , ———, and ———, 2012b: The role of institutions in the transformation of coping capacity to sustainable adaptive capacity. *Environ. Dev.*, **2**, 86–100, doi:10.1016/j.envdev.2012.03.017. <https://linkinghub.elsevier.com/retrieve/pii/S2211464512000565> (Accessed April 18, 2019).
- Bertram, C., and Coauthors, 2018: Targeted policies can compensate most of the increased sustainability risks in 1.5 °C mitigation scenarios. *Environ. Res. Lett.*, **13**, 64038. [Subject to Copy-editing](http://stacks.iop.org/1748-</p></div><div data-bbox=)

9326/13/i=6/a=064038.

- Bestelmeyer, B., and D. Briske, 2012: Grand challenges for resilience-based management of rangelands. *Rangel. Ecol. Manag.*, **65**, 654–663. <https://www.sciencedirect.com/science/article/pii/S1550742412501016> (Accessed November 10, 2018).
- Bestelmeyer, B. T., G. S. Okin, M. C. Duniway, S. R. Archer, N. F. Sayre, J. C. Williamson, and J. E. Herrick, 2015: Desertification, land use, and the transformation of global drylands. *Front. Ecol. Environ.*, **13**, 28–36, doi:10.1890/140162. <http://doi.wiley.com/10.1890/140162> (Accessed November 11, 2018).
- Beusen, A. H. W., A. F. Bouwman, L. P. H. Van Beek, J. M. Mogollón, J. J. Middelburg, and A. H. W. Beusen, 2016: Global riverine N and P transport to ocean increased during the 20th century despite increased retention along the aquatic continuum. *Biogeosciences*, **13**, 2441–2451, doi:10.5194/bg-13-2441-2016. www.biogeosciences.net/13/2441/2016/ (Accessed April 15, 2019).
- Bezner Kerr, R., 2013: Seed struggles and food sovereignty in northern Malawi. *J. Peasant Stud.*, **40**, 867–897, doi:10.1080/03066150.2013.848428. <http://www.tandfonline.com/doi/abs/10.1080/03066150.2013.848428> (Accessed April 15, 2019).
- Bhattacharjee, K., and B. Behera, 2017: Forest cover change and flood hazards in India. *Land use policy*, **67**, 436–448, doi:10.1016/j.landusepol.2017.06.013.
- Bhattacharyya, R., and Coauthors, 2015: Soil Degradation in India: Challenges and Potential Solutions. *Sustainability*, **7**, 3528–3570, doi:10.3390/su7043528. <http://www.mdpi.com/2071-1050/7/4/3528/>.
- Bhattachamishra, R., and C. B. Barrett, 2010: Community-based risk management arrangements: A Review. *World Dev.*, **38**, 923–932, doi:10.1016/j.worlddev.2009.12.017.
- Bidwell, D., T. Dietz, and D. Scavia, 2013: Fostering knowledge networks for climate adaptation. *Nat. Clim. Chang.*, **3**, 610–611, doi:10.1038/nclimate1931.
- Billen, G., L. Lassaletta, J. Garnier, J. L. Noë, E. Aguilera, and A. Sanz-Cobena, 2018: Opening to distant markets or local reconnection of agro-food systems? Environmental consequences at regional and global scales. *Farming Systems*, Elsevier.
- Binns, J. A., R. A. Maconachie, and A. I. Tanko, 2003: Water, land and health in urban and peri-urban food production: the case of Kano, Nigeria. *L. Degrad. Dev.*, **14**, 431–444, doi:10.1002/ldr.571. <http://doi.wiley.com/10.1002/ldr.571> (Accessed April 15, 2019).
- Binns, T., and K. Lynch, 1998: Feeding Africa’s growing cities into the 21st century: the potential of urban agriculture. *J. Int. Dev.*, **10**, 777–793, doi:10.1002/(SICI)1099-1328(199809)10:6<777::AID-JID532>3.0.CO;2-Z. <http://doi.wiley.com/10.1002/%28SICI%291099-1328%28199809%2910%3A6%3C777%3A%3AAID-JID532%3E3.0.CO%3B2-Z> (Accessed April 15, 2019).
- Birkmann, J., and Coauthors, 2015a: Scenarios for vulnerability: opportunities and constraints in the context of climate change and disaster risk. *Clim. Change*, **133**, 53–68, doi:10.1007/s10584-013-0913-2. <http://link.springer.com/10.1007/s10584-013-0913-2> (Accessed November 11, 2018).
- , and Coauthors, 2015b: Scenarios for vulnerability: opportunities and constraints in the context of climate change and disaster risk. *Clim. Change*, **133**, 53–68, doi:10.1007/s10584-013-0913-2. <http://link.springer.com/10.1007/s10584-013-0913-2> (Accessed August 11, 2018).
- Birthal, P. S., D. Roy, and D. S. Negi, 2015: Assessing the Impact of Crop Diversification on Farm Poverty in India. *World Dev.*, **72**, 70–92, doi:10.1016/J.WORLDDEV.2015.02.015.

- <https://www.sciencedirect.com/science/article/pii/S0305750X15000480> (Accessed May 31, 2018).
- Bisbis, M. B., N. Gruda, and M. Blanke, 2018: Potential impacts of climate change on vegetable production and product quality – A review. *J. Clean. Prod.*, **170**, 1602–1620, doi:10.1016/j.jclepro.2017.09.224.
- Bisht, I. ., P. . Mehta, K. . Negi, S. . Verma, R. . Tyagi, and S. . Garkoti, 2018: Farmers’ rights, local food systems, and sustainable household dietary diversification: A case of Uttarakhand Himalaya in north-western India. *Agroecol. Sustain. Food Syst.*, **42**, 77–113, doi:10.1080/21683565.2017.1363118. <https://www.tandfonline.com/doi/full/10.1080/21683565.2017.1363118> (Accessed June 1, 2018).
- Block, S., and P. Webb, 2001: The dynamics of livelihood diversification in post-famine Ethiopia. *Food Policy*, **26**, 333–350, doi:10.1016/S0306-9192(01)00015-X. <https://www.sciencedirect.com/science/article/abs/pii/S030691920100015X> (Accessed April 15, 2019).
- Bloom, J. D., and C. C. Hinrichs, 2011: Moving local food through conventional food system infrastructure: Value chain framework comparisons and insights. *Renew. Agric. Food Syst.*, **26**, 13–23, doi:10.1017/S1742170510000384. http://www.journals.cambridge.org/abstract_S1742170510000384 (Accessed April 15, 2019).
- Böcher, M., and M. Krott, 2014: The RIU model as an analytical framework for scientific knowledge transfer: the case of the “decision support system forest and climate change.” *Biodivers. Conserv.*, **23**, 3641–3656, doi:10.1007/s10531-014-0820-5.
- Bockstael, E., and F. Berkes, 2017: Using the capability approach to analyze contemporary environmental governance challenges in coastal Brazil. *Int. J. Commons*, **11**, 799–822, doi:10.18352/ijc.756. <https://www.thecommonsjournal.org/articles/10.18352/ijc.756> (Accessed November 10, 2018).
- Boisvenue, C., and S. W. Running, 2006: Impacts of climate change on natural forest productivity - evidence since the middle of the 20th century. *Glob. Chang. Biol.*, **12**, 862–882, doi:10.1111/j.1365-2486.2006.01134.x. <http://doi.wiley.com/10.1111/j.1365-2486.2006.01134.x> (Accessed April 4, 2019).
- de Bon, H., L. Parrot, and P. Moustier, 2010: Sustainable urban agriculture in developing countries. A review. *Agron. Sustain. Dev.*, **30**, 21–32, doi:10.1051/agro:2008062. <http://link.springer.com/10.1051/agro:2008062> (Accessed April 15, 2019).
- Bonan, G. B., 2008a: Forests and climate change: forcings, feedbacks, and the climate benefits of forests. *Science*, **320**, 1444–1449, doi:10.1126/science.1155121. <http://www.ncbi.nlm.nih.gov/pubmed/18556546> (Accessed April 4, 2019).
- Bonan, G. B., 2008b: Forests and climate change: Forcings, feedbacks, and the climate benefits of forests. *Science (80-.)*, **320**, 1444–1449, doi:10.1126/science.1155121.
- Bonn, A., M. Reed, C. Evans, H. Joosten, and C. B. Services, 2014: Investing in nature: developing ecosystem service markets for peatland restoration. *Ecosystem*, **9**, 54–65. <http://www.sciencedirect.com/science/article/pii/S2212041614000692> (Accessed December 20, 2017).
- Bonny, S., 1993: Is agriculture using more and more energy? A French case study. *Agric. Syst.*, **43**, 51–66, doi:10.1016/0308-521X(93)90092-G. <https://www.sciencedirect.com/science/article/pii/0308521X9390092G> (Accessed April 15, 2019).
- Bonsch, M., A. Popp, A. Biewald, ... S. R.-G. E., and undefined 2015, Environmental flow provision: Implications for agricultural water and land-use at the global scale. *Elsevier*.

- Bonsch, M., and Coauthors, 2015a: Environmental flow provision: Implications for agricultural water and land-use at the global scale. *Glob. Environ. Chang.*, **30**, 113–132, doi:10.1016/j.gloenvcha.2014.10.015.
- , and Coauthors, 2015b: Environmental flow provision: Implications for agricultural water and land-use at the global scale. *Glob. Environ. Chang.*, **30**, 113–132, doi:10.1016/j.gloenvcha.2014.10.015. <https://www.sciencedirect.com/science/article/pii/S0959378014001964> (Accessed November 11, 2018).
- , and Coauthors, 2016: Trade-offs between land and water requirements for large-scale bioenergy production. *GCB Bioenergy*, **8**, 11–24, doi:10.1111/gcbb.12226. <http://doi.wiley.com/10.1111/gcbb.12226> (Accessed May 30, 2018).
- Boone, R. B., R. T. Conant, J. Sircely, P. K. Thornton, and M. Herrero, 2018: Climate change impacts on selected global rangeland ecosystem services. *Glob. Chang. Biol.*, **24**, 1382–1393, doi:10.1111/gcb.13995. <https://onlinelibrary.wiley.com/doi/abs/10.1111/gcb.13995> (Accessed April 4, 2019).
- Börjesson, P., and L. Gustavsson, 2000: Greenhouse gas balances in building construction: wood versus concrete from life-cycle and forest land-use perspectives. *Energy Policy*, **28**, 575–588, doi:10.1016/S0301-4215(00)00049-5. <https://www.sciencedirect.com/science/article/pii/S0301421500000495> (Accessed April 15, 2019).
- Borrelli, P., and Coauthors, 2017: An assessment of the global impact of 21st century land use change on soil erosion. *Nat. Commun.*, **8**, 2013, doi:10.1038/s41467-017-02142-7. <http://www.nature.com/articles/s41467-017-02142-7> (Accessed April 5, 2019).
- Bourgoin, J., 2012: Sharpening the understanding of socio-ecological landscapes in participatory land-use planning. A case study in Lao PDR. *Appl. Geogr.*, **34**, 99–110, doi:10.1016/j.apgeog.2011.11.003. <http://linkinghub.elsevier.com/retrieve/pii/S0143622811002244> (Accessed November 10, 2018).
- Bouwer, L. M., E. Papyrakis, J. Poussin, C. Pfuerscheller, and A. H. Thieken, 2014: The Costing of Measures for Natural Hazard Mitigation in Europe. *Nat. Hazards Rev.*, **15**, 4014010, doi:10.1061/(ASCE)NH.1527-6996.0000133. <http://ascelibrary.org/doi/10.1061/%28ASCE%29NH.1527-6996.0000133> (Accessed June 1, 2018).
- Bowman, A., 2015: Sovereignty, Risk and Biotechnology: Zambia’s 2002 GM Controversy in Retrospect. *Dev. Change*, **46**, 1369–1391, doi:10.1111/dech.12196. <http://doi.wiley.com/10.1111/dech.12196> (Accessed June 1, 2018).
- Bowman, M. S., and D. Zilberman, 2013: Economic Factors Affecting Diversified Farming Systems. *Ecol. Soc.*, **18**, art33, doi:10.5751/ES-05574-180133. <http://www.ecologyandsociety.org/vol18/iss1/art33/> (Accessed April 15, 2019).
- Boyd, J., and S. Banzhaf, 2007: What are ecosystem services? The need for standardized environmental accounting units. *Ecol. Econ.*, **63**, 616–626, doi:10.1016/J.ECOLECON.2007.01.002. <https://www.sciencedirect.com/science/article/pii/S0921800907000341> (Accessed April 4, 2019).
- Boysen, L. R., W. Lucht, and D. Gerten, 2017a: Trade-offs for food production, nature conservation and climate limit the terrestrial carbon dioxide removal potential. *Glob. Chang. Biol.*, **23**, 4303–4317, doi:10.1111/gcb.13745.
- , ——, and ——, 2017b: Trade-offs for food production, nature conservation and climate limit the terrestrial carbon dioxide removal potential. *Glob. Chang. Biol.*, **23**, 4303–4317, doi:10.1111/gcb.13745. <http://doi.wiley.com/10.1111/gcb.13745> (Accessed April 15, 2019).
- Braat, L., 2018: Five reasons why the Science publication “Assessing nature’s contributions to people” (Diaz et al. 2018) would not have been accepted in Ecosystem Services. *Ecosyst. Serv.*, **30**, A1–A2.

- Bradford, K. J., P. Dahal, J. Van Asbrouck, K. Kunusoth, P. Bello, J. Thompson, and F. Wu, 2018: The dry chain: Reducing postharvest losses and improving food safety in humid climates. *Trends Food Sci. Technol.*, **71**, 84–93, doi:10.1016/j.tifs.2017.11.002. <https://www.sciencedirect.com/science/article/pii/S092422441730482X> (Accessed November 8, 2018).
- Bradshaw, B., H. Dolan, and B. Smit, 2004: Farm-Level Adaptation to Climatic Variability and Change: Crop Diversification in the Canadian Prairies. *Clim. Change*, **67**, 119–141, doi:10.1007/s10584-004-0710-z. <http://link.springer.com/10.1007/s10584-004-0710-z> (Accessed April 4, 2019).
- Braun, J. von, B. Algieri, M. K.-W. F. Policy, and U. 2014, 2014: World food system disruptions in the early 2000s: causes, impacts and cures. *World Food Policy*, **1**, 34–55. https://www.researchgate.net/profile/Matthias_Kalkuhl/publication/283391085_World_Food_System_Disruptions_in_the_Early_2000s_Causes_Impacts_and_Cures/links/57192bd508ae986b8b7b3208.pdf (Accessed November 11, 2018).
- de Brauw, A., 2011: Migration and child development during the food price crisis in El Salvador. *Food Policy*, **36**, 28–40, doi:10.1016/J.FOODPOL.2010.11.002. <https://www.sciencedirect.com/science/article/abs/pii/S0306919210001168> (Accessed April 15, 2019).
- Bren d'Amour, C., and Coauthors, 2016: Future urban land expansion and implications for global croplands. *Proc. Natl. Acad. Sci.*, 201606036, doi:10.1073/pnas.1606036114. <http://www.pnas.org/content/114/34/8939.short> (Accessed May 1, 2018).
- Brenkert, A. L., and E. L. Malone, 2005: Modeling vulnerability and resilience to climate change: A case study of India and Indian states. *Clim. Change*, **72**, 57–102, doi:10.1007/s10584-005-5930-3.
- Brew-Hammond, A., 2010: Energy access in Africa: Challenges ahead. *Energy Policy*, **38**, 2291–2301, doi:10.1016/J.ENPOL.2009.12.016. <https://www.sciencedirect.com/science/article/pii/S0301421509009707> (Accessed April 15, 2019).
- Briber, B. M., L. R. Hutryra, A. B. Reinmann, S. M. Raciti, V. K. Dearborn, C. E. Holden, and A. L. Dunn, 2015: Tree Productivity Enhanced with Conversion from Forest to Urban Land Covers. *PLoS One*, **10**, e0136237, doi:10.1371/journal.pone.0136237. <http://dx.plos.org/10.1371/journal.pone.0136237> (Accessed May 31, 2018).
- Brindha, K., and P. Pavelic, 2016: Identifying priority watersheds to mitigate flood and drought impacts by novel conjunctive water use management. *Mitig. Adapt. Strateg. Glob. Chang.*, **75**, 399, doi:10.1007/s12665-015-4989-z.
- Brink, B. J. E. ten., and Coauthors, 2018: Chapter 7: Scenarios of IPBES, land degradation and restoration. *Land: The IPBES assessment report on degradation and restoration*, L. Montanarella, R. Scholes, and A. Brainich, Eds., Intergovernmental Ecosystem, Platform on Biodiversity and Services, Bonn, Germany, 531–589.
- Brinkley, C., E. Birch, and A. Keating, 2013: Feeding cities: charting a research and practice agenda towards food security. *J. Agric. food Syst. community Dev.*, **3**, 81–87. <http://www.foodsystemsjournal.org/index.php/fsj/article/view/190> (Accessed May 1, 2018).
- Brinkley, C., E. Birch, and A. Keating, 2016: Feeding cities: Charting a research and practice agenda toward food security. *Agric. Food Secur.*, **3**, 81–87. <http://www.foodsystemsjournal.org/index.php/fsj/article/view/190> (Accessed May 1, 2018).
- Brinkman, H., S. De Pee, I. Sanogo, L. Subran, and M. Bloem, 2009: High Food Prices and the Global Financial Crisis Have Reduced Access to Nutritious Food and Worsened Nutritional Status and Health. *J. Nutr.*, **140**, 153S–161S. <https://academic.oup.com/jn/article-abstract/140/1/153S/4600303>

(Accessed November 11, 2018).

- Brinkman, H.-J., S. de Pee, I. Sanogo, L. Subran, and M. W. Bloem, 2010: High Food Prices and the Global Financial Crisis Have Reduced Access to Nutritious Food and Worsened Nutritional Status and Health. *J. Nutr.*, **140**, 153S–161S, doi:10.3945/jn.109.110767. <https://academic.oup.com/jn/article/140/1/153S/4600303> (Accessed April 15, 2019).
- Briske, D. D., L. A. Joyce, H. W. Polley, J. R. Brown, K. Wolter, J. A. Morgan, B. A. McCarl, and D. W. Bailey, 2015: Climate-change adaptation on rangelands: linking regional exposure with diverse adaptive capacity. *Front. Ecol. Environ.*, **13**, 249–256, doi:10.1890/140266. <http://doi.wiley.com/10.1890/140266> (Accessed November 11, 2018).
- Brondizio, E. S., E. Ostrom, and O. R. Young, 2009: Connectivity and the Governance of Multilevel Social-Ecological Systems: The Role of Social Capital. *Annu. Rev. Environ. Resour.*, **34**, 253–278, doi:10.1146/annurev.environ.020708.100707. <http://www.annualreviews.org/doi/10.1146/annurev.environ.020708.100707> (Accessed November 10, 2018).
- Brooks, N., W. N. Adger, and P. M. Kelly, 2005: The determinants of vulnerability and adaptive capacity at the national level and the implications for adaptation. *Glob. Environ. Chang.*, **15**, 151–163, doi:10.1016/j.gloenvcha.2004.12.006.
- Brueckner, J. K., 2000: Urban Sprawl: Diagnosis and Remedies. *Int. Reg. Sci. Rev.*, **23**, 160–171, doi:10.1177/016001700761012710. <http://journals.sagepub.com/doi/10.1177/016001700761012710> (Accessed April 15, 2019).
- Brundu, G., and D. M. Richardson, 2016: Planted forests and invasive alien trees in Europe: A Code for managing existing and future plantings to mitigate the risk of negative impacts from invasions. *NeoBiota*, **30**, 5–47, doi:10.3897/neobiota.30.7015. <http://neobiota.pensoft.net/articles.php?id=7015> (Accessed November 8, 2018).
- Bryan, E., T. T. Deressa, G. A. Gbetibouo, and C. Ringler, 2009a: Adaptation to climate change in Ethiopia and South Africa: options and constraints. *Environ. Sci. Policy*, **12**, 413–426, doi:10.1016/J.ENVSCI.2008.11.002. <https://www.sciencedirect.com/science/article/pii/S1462901108001263> (Accessed June 4, 2018).
- , ———, ———, and ———, 2009b: Adaptation to climate change in Ethiopia and South Africa: options and constraints. *Environ. Sci. Policy*, **12**, 413–426, doi:10.1016/j.envsci.2008.11.002. <http://linkinghub.elsevier.com/retrieve/pii/S1462901108001263> (Accessed November 7, 2018).
- Bryan, E., C. Ringler, B. Okoba, C. Roncoli, S. Silvestri, and M. Herrero, 2013: Adapting agriculture to climate change in Kenya: Household strategies and determinants. *J. Environ. Manage.*, **114**, 26–35. <https://www.sciencedirect.com/science/article/pii/S0301479712005415> (Accessed November 11, 2018).
- Bryceson, D. F., 1999: African rural labour, income diversification & livelihood approaches: a long-term development perspective. *Rev. Afr. Polit. Econ.*, **26**, 171–189, doi:10.1080/03056249908704377. <http://www.tandfonline.com/doi/abs/10.1080/03056249908704377> (Accessed November 11, 2018).
- Burney, J., L. Woltering, ... M. B.-P. of the, and undefined 2010, Solar-powered drip irrigation enhances food security in the Sudano-Sahel. *Natl. Acad. Sci.*, <http://www.pnas.org/content/107/5/1848.short> (Accessed November 8, 2018).
- Burney, J., L. Woltering, M. Burke, R. Naylor, and D. Pasternak, 2010: Solar-powered drip irrigation enhances food security in the Sudano-Sahel. *Proc. Natl. Acad. Sci.*, **107**, 1848–1853, doi:10.1073/pnas.0909678107.

- Burns, W., and S. Nicholson, 2017: Bioenergy and carbon capture with storage (BECCS): the prospects and challenges of an emerging climate policy response. *J. Environ. Stud. Sci.*, **7**, 527–534, doi:10.1007/s13412-017-0445-6. <http://link.springer.com/10.1007/s13412-017-0445-6> (Accessed April 15, 2019).
- Busse, H., W. Jogo, G. Levenson, F. Asfaw, and H. Tesfay, 2017: Prevalence and predictors of stunting and underweight among children under 5 years in Tigray, Ethiopia: Implications for nutrition-sensitive agricultural interventions. *J. Hunger Environ. Nutr.*, 1–20, doi:10.1080/19320248.2017.1393364. <https://www.tandfonline.com/doi/full/10.1080/19320248.2017.1393364> (Accessed June 2, 2018).
- Bustamante, M., and Coauthors, 2014a: Co-benefits, trade-offs, barriers and policies for greenhouse gas mitigation in the agriculture, forestry and other land use (AFOLU) sector. *Glob. Chang. Biol.*, **20**, 3270–3290, doi:10.1111/gcb.12591. <http://doi.wiley.com/10.1111/gcb.12591> (Accessed December 18, 2017).
- Bustamante, M., and Coauthors, 2014b: Co-benefits, trade-offs, barriers and policies for greenhouse gas mitigation in the agriculture, forestry and other land use (AFOLU) sector. *Glob. Chang. Biol.*, **20**, 3270–3290, doi:10.1111/gcb.12591. <http://doi.wiley.com/10.1111/gcb.12591> (Accessed May 30, 2018).
- , and Coauthors, 2014c: Co-benefits, trade-offs, barriers and policies for greenhouse gas mitigation in the agriculture, forestry and other land use (AFOLU) sector. *Glob. Chang. Biol.*, **20**, 3270–3290, doi:10.1111/gcb.12591. <http://doi.wiley.com/10.1111/gcb.12591> (Accessed December 18, 2017).
- Byerlee, D., T. Jayne, and R. Myers, 2006: Managing food price risks and instability in a liberalizing market environment: Overview and policy options. *Food Policy*, **31**, 275–287. <https://www.sciencedirect.com/science/article/pii/S030691920600025X> (Accessed November 11, 2018).
- Byers, E., M. Gidden, and D. Lecl, 2018: Global exposure and vulnerability to multi-sector development and climate change hotspots Global exposure and vulnerability to multi-sector development and climate change hotspots. *Environ. Res. Lett.*,
- Cacho, J. F., M. C. Negri, C. R. Zumpf, and P. Campbell, 2018: Introducing perennial biomass crops into agricultural landscapes to address water quality challenges and provide other environmental services. *Wiley Interdiscip. Rev. Energy Environ.*, **7**, e275.
- Cai, H., X. Yang, and X. Xu, 2013: Spatiotemporal Patterns of Urban Encroachment on Cropland and Its Impacts on Potential Agricultural Productivity in China. *Remote Sens.*, **5**, 6443–6460, doi:10.3390/rs5126443. <http://www.mdpi.com/2072-4292/5/12/6443> (Accessed May 31, 2018).
- Calder, I. R., 2005: *Blue Revolution: Integrated Land and Water Resource Management - Ian R. Calder - Google Books*. Second. Earthscan, London, 347 pp. [https://books.google.com/jm/books?hl=en&lr=&id=fnUjXtHjJcC&oi=fnd&pg=PR2&dq=Calder,+I.R.+\(2005\)+Blue+revolution:+integrated+land+and+water+resource+management.+Routledge.&ots=IU3nyCbdAu&sig=qAuizNG8xANuYdSUgmEZPbbDapg&redir_esc=y#v=onepage&q=Calder%2C](https://books.google.com/jm/books?hl=en&lr=&id=fnUjXtHjJcC&oi=fnd&pg=PR2&dq=Calder,+I.R.+(2005)+Blue+revolution:+integrated+land+and+water+resource+management.+Routledge.&ots=IU3nyCbdAu&sig=qAuizNG8xANuYdSUgmEZPbbDapg&redir_esc=y#v=onepage&q=Calder%2C) (Accessed December 22, 2017).
- Calow, R. C., A. M. MacDonald, A. L. Nicol, and N. S. Robins, 2010: Ground Water Security and Drought in Africa: Linking Availability, Access, and Demand. *Ground Water*, **48**, 246–256, doi:10.1111/j.1745-6584.2009.00558.x. <http://doi.wiley.com/10.1111/j.1745-6584.2009.00558.x> (Accessed April 15, 2019).
- Calvin, K., M. Wise, L. Clarke, J. Edmonds, P. Kyle, P. Luckow, and A. Thomson, 2013a: Implications of simultaneously mitigating and adapting to climate change: Initial experiments using GCAM. *Clim. Change*, **117**, 545–560, doi:10.1007/s10584-012-0650-y.

- , ——, ——, ——, ——, ——, and ——, 2013b: Implications of simultaneously mitigating and adapting to climate change: Initial experiments using GCAM. *Clim. Change*, **117**, 545–560, doi:10.1007/s10584-012-0650-y. <http://link.springer.com/10.1007/s10584-012-0650-y> (Accessed May 30, 2018).
- Calvin, K., M. Wise, P. Kyle, P. Patel, L. Clarke, and J. Edmonds, 2014: Trade-offs of different land and bioenergy policies on the path to achieving climate targets. *Clim. Change*, **123**, 691–704, doi:10.1007/s10584-013-0897-y.
- Calvin, K., M. Wise, P. Luckow, P. Kyle, L. Clarke, and J. Edmonds, 2016a: Implications of uncertain future fossil energy resources on bioenergy use and terrestrial carbon emissions. *Clim. Change*, **136**, doi:10.1007/s10584-013-0923-0.
- Calvin, K., and Coauthors, 2017a: The SSP4: A world of deepening inequality. *Glob. Environ. Chang.*, **42**, 284–296, doi:10.1016/j.gloenvcha.2016.06.010.
- , and Coauthors, 2017b: The SSP4: A world of deepening inequality. *Glob. Environ. Chang.*, **42**, 284–296, doi:10.1016/j.gloenvcha.2016.06.010. <https://linkinghub.elsevier.com/retrieve/pii/S095937801630084X> (Accessed November 10, 2018).
- , and Coauthors, 2017c: The SSP4: A world of deepening inequality. *Glob. Environ. Chang.*, **42**, 284–296, doi:10.1016/j.gloenvcha.2016.06.010. <http://dx.doi.org/10.1016/j.gloenvcha.2016.06.010>.
- , and Coauthors, 2019: GCAM v5.1: representing the linkages between energy, water, land, climate, and economic systems. *Geosci. Model Dev.*, **12**, 677–698, doi:10.5194/gmd-12-677-2019. <https://www.geosci-model-dev.net/12/677/2019/> (Accessed April 14, 2019).
- Calvin, K. V., R. Beach, A. Gurgel, M. Labriet, and A. M. Loboguerrero Rodriguez, 2016b: Agriculture, forestry, and other land-use emissions in Latin America. *Energy Econ.*, **56**, 615–624, doi:<https://doi.org/10.1016/j.eneco.2015.03.020>. <http://www.sciencedirect.com/science/article/pii/S0140988315001127>.
- Campbell, B. C., and J. R. Veteto, 2015: Free seeds and food sovereignty: anthropology and grassroots agrobiodiversity conservation strategies in the US South. *J. Polit. Ecol.*, **22**. http://jpe.library.arizona.edu/volume_22/CampbellVeteto.pdf (Accessed May 31, 2018).
- Campbell, B. M., P. Thornton, R. Zougmore, P. van Asten, and L. Lipper, 2014: Sustainable intensification: What is its role in climate smart agriculture? *Curr. Opin. Environ. Sustain.*, **8**, 39–43, doi:10.1016/J.COSUST.2014.07.002. <https://www.sciencedirect.com/science/article/pii/S1877343514000359> (Accessed May 30, 2018).
- Campbell, J. E., D. B. Lobell, R. C. Genova, and C. B. Field, 2008: The global potential of bioenergy on abandoned agriculture lands. *Environ. Sci. Technol.*, **42**, 5791–5794, doi:10.1021/es800052w.
- Campbell, J. R., 2015: Development, global change and traditional food security in Pacific Island countries. *Reg. Environ. Chang.*, **15**, 1313–1324, doi:10.1007/s10113-014-0697-6. <http://link.springer.com/10.1007/s10113-014-0697-6> (Accessed May 31, 2018).
- Caplow, S., P. Jagger, K. Lawlor, and E. Sills, 2011: Evaluating land use and livelihood impacts of early forest carbon projects: Lessons for learning about REDD+. *Environ. Sci. Policy*, **14**, 152–167, doi:10.1016/j.envsci.2010.10.003.
- Capone, R., H. El Bilali, P. Debs, G. Cardone, and N. Driouech, 2014: Food System Sustainability and Food Security: Connecting the Dots. *J. Food Secur.*, **2**, 13–22, doi:10.12691/JFS-2-1-2.
- Cardoso, A. S., A. Berndt, A. Leytem, B. J. R. Alves, I. das N. O. de Carvalho, L. H. de Barros Soares, S. Urquiaga, and R. M. Boddey, 2016: Impact of the intensification of beef production in Brazil on greenhouse gas emissions and land use. *Agric. Syst.*, **143**, 86–96, doi:10.1016/J.AGSY.2015.12.007.

- <https://www.sciencedirect.com/science/article/pii/S0308521X15300652> (Accessed June 2, 2018).
- Carlson, K. M., and L. M. Curran, 2013: Refined carbon accounting for oil palm agriculture: disentangling potential contributions of indirect emissions and smallholder farmers. *Carbon Manag.*, **4**, 347–349, doi:10.4155/cmt.13.39. <http://www.tandfonline.com/doi/abs/10.4155/cmt.13.39> (Accessed April 12, 2019).
- , and Coauthors, 2018: Effect of oil palm sustainability certification on deforestation and fire in Indonesia. *Proc. Natl. Acad. Sci. U. S. A.*, **115**, 121–126, doi:10.1073/pnas.1704728114. <http://www.ncbi.nlm.nih.gov/pubmed/29229857> (Accessed June 2, 2018).
- Carmenta, R., A. Zabala, W. Daeli, J. P.-G. E. Change, and undefined 2017, Perceptions across scales of governance and the Indonesian peatland fires. *Elsevier*,.
- Carpenter, S. R., and Coauthors, 2009: Science for managing ecosystem services: Beyond the Millennium Ecosystem Assessment. *Proc. Natl. Acad. Sci. U. S. A.*, **106**, 1305–1312, doi:10.1073/pnas.0808772106. <http://www.ncbi.nlm.nih.gov/pubmed/19179280> (Accessed April 18, 2019).
- Carreño, M. L., O. D. Cardona, and A. H. Barbat, 2007: A disaster risk management performance index. *Nat. Hazards*, **41**, 1–20, doi:10.1007/s11069-006-9008-y. <https://doi.org/10.1007/s11069-006-9008-y>.
- Carter, D. R., R. T. Fahey, K. Dreisilker, M. B. Bialecki, and M. L. Bowles, 2015: Assessing patterns of oak regeneration and C storage in relation to restoration-focused management, historical land use, and potential trade-offs. *For. Ecol. Manage.*, **343**, 53–62, doi:<https://doi.org/10.1016/j.foreco.2015.01.027>.
- Carvalho, J. L. N., T. W. Hudiburg, H. C. J. Franco, and E. H. Delucia, 2016: Contribution of above- and belowground bioenergy crop residues to soil carbon. *GCB Bioenergy*, doi:10.1111/gcbb.12411.
- Cava, M. G. B., N. A. L. Pilon, M. C. Ribeiro, and G. Durigan, 2018: Abandoned pastures cannot spontaneously recover the attributes of old-growth savannas. *J. Appl. Ecol.*, **55**, 1164–1172, doi:10.1111/1365-2664.13046. <http://doi.wiley.com/10.1111/1365-2664.13046> (Accessed June 2, 2018).
- Cavanagh, C., A. Chemarum, P. Vedeld, and J. Petursson, 2017: Old wine, new bottles? Investigating the differential adoption of “climate-smart” agricultural practices in western Kenya. *J. Rural Stud.*, **56**, 114–123. <https://www.sciencedirect.com/science/article/pii/S0743016716304697> (Accessed November 10, 2018).
- Cerri, C. E. P., C. C. Cerri, S. M. F. Maia, M. R. Cherubin, B. J. Feigl, and R. Lal, 2018: Reducing Amazon Deforestation through Agricultural Intensification in the Cerrado for Advancing Food Security and Mitigating Climate Change. *Sustainability*, **10**, 989, doi:10.3390/su10040989.
- CGIAR, 2016: The drought crisis in the Central Highlands of Vietnam. 1–36.
- Chaboud, G., and B. Daviron, 2017: Food losses and waste: navigating the inconsistencies. *Glob. Food Sec.*, **12**, 1–7. <https://www.sciencedirect.com/science/article/pii/S2211912416300499> (Accessed May 10, 2018).
- Chadwick, D. R., and Coauthors, 2014: Optimizing chamber methods for measuring nitrous oxide emissions from plot-based agricultural experiments. *Eur. J. Soil Sci.*, doi:10.1111/ejss.12117.
- Chakauya, E., G. Beyene, and R. K. Chikwamba, 2009: *South African journal of science*. Academy of Science of South Africa, 174-181 pp. http://www.scielo.org.za/scielo.php?script=sci_arttext&pid=S0038-23532009000300010 (Accessed April 15, 2019).

- Challinor, A. J., J. Watson, D. B. Lobell, S. M. Howden, D. R. Smith, and N. Chhetri, 2014: A meta-analysis of crop yield under climate change and adaptation Feeding a growing global population in a changing climate presents a significant challenge to society. *Nat. Clim. Chang.*, **4**, 287–291, doi:10.1038/NCLIMATE2153.
- Chamen, W., A. P. Moxey, W. Towers, B. Balana, and P. D. Hallett, 2015: Mitigating arable soil compaction: A review and analysis of available cost and benefit data. *Soil Tillage Res.*, **146**, 10–25, doi:10.1016/J.STILL.2014.09.011.
<https://www.sciencedirect.com/science/article/pii/S0167198714001901> (Accessed May 31, 2018).
- Chang, Y., G. Li, Y. Yao, L. Zhang, and C. Yu, 2016: Quantifying the water-energy-food nexus: Current status and trends. *Energies*, doi:10.3390/en9020065.
- Chappell, M. J., J. R. Moore, and A. A. Heckelman, 2016: Participation in a city food security program may be linked to higher ant alpha- and beta-diversity: an exploratory case from Belo Horizonte, Brazil. *Agroecol. Sustain. Food Syst.*, **40**, 804–829, doi:10.1080/21683565.2016.1160020.
<https://www.tandfonline.com/doi/full/10.1080/21683565.2016.1160020> (Accessed May 1, 2018).
- Chaturvedi, V., M. Hejazi, J. Edmonds, L. Clarke, P. Kyle, E. Davies, and M. Wise, 2013: Climate mitigation policy implications for global irrigation water demand. *Mitig. Adapt. Strateg. Glob. Chang.*, **20**, 389–407, doi:10.1007/s11027-013-9497-4.
- Cheesman, S., C. Thierfelder, N. S. Eash, G. T. Kassie, and E. Frossard, 2016: Soil carbon stocks in conservation agriculture systems of Southern Africa. *Soil Tillage Res.*, **156**, 99–109, doi:10.1016/J.STILL.2015.09.018.
<https://www.sciencedirect.com/science/article/pii/S0167198715300350> (Accessed June 5, 2018).
- Chen, H., B. Jia, and S. S. Y. Lau, 2008: Sustainable urban form for Chinese compact cities: Challenges of a rapid urbanized economy. *Habitat Int.*, **32**, 28–40, doi:10.1016/J.HABITATINT.2007.06.005.
<https://www.sciencedirect.com/science/article/pii/S0197397507000367> (Accessed April 15, 2019).
- Chen, J., 2007a: Rapid urbanization in China: A real challenge to soil protection and food security. *Catena*, **69**, 1–15, doi:10.1016/J.CATENA.2006.04.019.
<https://www.sciencedirect.com/science/article/pii/S0341816206000920> (Accessed May 31, 2018).
- , 2007b: Rapid urbanization in China: A real challenge to soil protection and food security. *Catena*, **69**, 1–15, doi:10.1016/J.CATENA.2006.04.019.
<https://www.sciencedirect.com/science/article/pii/S0341816206000920> (Accessed November 9, 2018).
- Chen, L., J. Wang, W. Wei, B. Fu, and D. Wu, 2010: Effects of landscape restoration on soil water storage and water use in the Loess Plateau Region, China. *For. Ecol. Manage.*, **259**, 1291–1298, doi:10.1016/J.FORECO.2009.10.025.
<https://www.sciencedirect.com/science/article/pii/S0378112709007671> (Accessed June 5, 2018).
- Chen, W., 2017: Environmental externalities of urban river pollution and restoration: A hedonic analysis in Guangzhou (China). *Landsc. Urban Plan.*, **157**, 170–179.
<https://www.sciencedirect.com/science/article/pii/S0169204616301177> (Accessed May 1, 2018).
- Cherubini, F., S. Vezhapparambu, W. Bogren, R. Astrup, and A. H. Strømman, 2017: Spatial, seasonal, and topographical patterns of surface albedo in Norwegian forests and cropland. *Int. J. Remote Sens.*, **38**, 4565–4586, doi:10.1080/01431161.2017.1320442.
- Chhatre, A., and A. Agrawal, 2009: Trade-offs and synergies between carbon storage and livelihood benefits from forest commons. *Proc. Natl. Acad. Sci.*, **106**, 17667–17670, doi:10.1073/pnas.0905308106.
<http://www.pnas.org/content/106/42/17667%7B%25%7D5Cnhttp://www.ncbi.nlm.nih.gov/pubmed/>

- 19815522%7B%25%7D5Cnhttp://www.pnas.org/content/106/42/17667.full.pdf%7B%25%7D5Cnhttp://www.pnas.org/content/106/42/17667.short.
- Chow, J., 2018: Mangrove management for climate change adaptation and sustainable development in coastal zones. *J. Sustain. For.*, **37**, 139–156, doi:10.1080/10549811.2017.1339615.
- Chum, H., and Coauthors, 2011: Bioenergy. *IPCC Special Report on Renewable Energy Sources and Climate Change Mitigation*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- Chum, H., and Coauthors, 2013: Bioenergy. *IPCC Special Report on Renewable Energy Sources and Climate Change Mitigation.*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge.
- Ciais, P., and Coauthors, 2013: Carbon and Other Biogeochemical Cycles. *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, Cambridge University Press, 465–570.
- Ciccarese, L., A. Mattsson, and D. Pettenella, 2012: Ecosystem services from forest restoration: thinking ahead. *New For.*, **43**, 543–560, doi:10.1007/s11056-012-9350-8.
- Claassen, R., F. Carriazo, J. Cooper, D. Hellerstein, and K. Ueda, 2011: *Grassland to Cropland Conversion in the Northern Plains. The Role of Crop Insurance, Commodity, and Disaster Programs*. 85 pp.
- Clark, M., and D. Tilman, 2017: Comparative analysis of environmental impacts of agricultural production systems, agricultural input efficiency, and food choice. *Environ. Res. Lett.*, **12**, 64016, doi:10.1088/1748-9326/aa6cd5. <http://stacks.iop.org/1748-9326/12/i=6/a=064016?key=crossref.f80d1b1b72259fb25e1f060a8362e9ca> (Accessed May 30, 2018).
- Clarke, L., and K. Jiang, 2014a: Assessing Transformation Pathways Climate Change 2013: Mitigation of Climate Change Contribution of Working Group 3rd to the 5th Assessment Report. https://scholar.google.com/scholar?hl=en&as_sdt=0%2C5&q=Clarke%2C+L.+E.%2C+and+Coauthors%2C+2014%3A+Assessing+transformation+pathways.+Climate+Change+2014%3A+Mitigation+of+Climate+Change.+Contribution+of+Working+Group+III+to+the+Fifth+Assessment+Report+of+the+Intergovernmental+Panel+on+Climate+Change%2C&btnG= (Accessed July 4, 2018).
- , and ———, 2014b: Assessing Transformation Pathways Climate Change 2013: Mitigation of Climate Change Contribution of Working Group 3rd to the 5th Assessment Report. https://scholar.google.com/scholar?hl=en&as_sdt=0%2C5&q=Clarke%2C+L.+E.%2C+and+Coauthors%2C+2014%3A+Assessing+transformation+pathways.+Climate+Change+2014%3A+Mitigation+of+Climate+Change.+Contribution+of+Working+Group+III+to+the+Fifth+Assessment+Report+of+the+Intergovernmental+Panel+on+Climate+Change%2C+413-510+&btnG= (Accessed November 11, 2018).
- Clarke, L. E., and Coauthors, 2014: Assessing transformation pathways. *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, 413–510 <http://scholar.google.com/scholar?hl=en&btnG=Search&q=intitle:Assessing+Transformation+Pathways#4>.
- Claudio, L., 2012: Our Food: Packaging & Public Health. *Environ. Health Perspect.*, **120**, doi:10.1289/ehp.120-a232. <https://ehp.niehs.nih.gov/doi/10.1289/ehp.120-a232> (Accessed April 15, 2019).
- Coakley, J., Atmospheric physics: Reflections on aerosol cooling. *nature.com.*, <https://www.nature.com/articles/4381091a> (Accessed November 11, 2018).

- Cohen, M. J., and J. L. Garrett, 2010: The food price crisis and urban food (in)security. *Environ. Urban.*, **22**, 467–482, doi:10.1177/0956247810380375. <http://journals.sagepub.com/doi/10.1177/0956247810380375> (Accessed April 15, 2019).
- Cohen, M. J., J. Clapp, and Centre for International Governance Innovation., 2009: *The global food crisis: governance challenges and opportunities*. Wilfrid Laurier University Press, 267 pp. https://books.google.co.nz/books?hl=en&lr=&id=Wa_wKWgbTSgC&oi=fnd&pg=PR5&dq=The+Global+Food+Crisis:+Governance+Challenges+and+Opportunities.+&ots=yWgJADp8R_&sig=w6fa uWwuIuQyUQgrtVGWp_0aNAU#v=onepage&q=The+Global+Food+Crisis%3A+Governance+Challenges+and+Opportunities.&f=false (Accessed June 1, 2018).
- Cohn, A. S., J. Gil, T. Berger, H. Pellegrina, and C. Toledo, 2016: Patterns and processes of pasture to crop conversion in Brazil: Evidence from Mato Grosso State. *Land use policy*, **55**, 108–120, doi:10.1016/J.LANDUSEPOL.2016.03.005. <https://www.sciencedirect.com/science/article/pii/S0264837716301909> (Accessed June 2, 2018).
- , P. Newton, J. D. B. Gil, L. Kuhl, L. Samberg, V. Ricciardi, J. R. Manly, and S. Northrop, 2017: Smallholder Agriculture and Climate Change. *Annu. Rev. Environ. Resour.*, **42**, 347–375, doi:10.1146/annurev-environ-102016-060946. <http://www.annualreviews.org/doi/10.1146/annurev-environ-102016-060946> (Accessed November 11, 2018).
- Coley, D., M. Howard, and M. Winter, 2009: Local food, food miles and carbon emissions: A comparison of farm shop and mass distribution approaches. *Food Policy*, **34**, 150–155, doi:10.1016/J.FOODPOL.2008.11.001. <https://www.sciencedirect.com/science/article/abs/pii/S0306919208000997> (Accessed April 15, 2019).
- Collins, K., and R. Ison, 2009: Jumping off Arnstein’s ladder: Social learning as a new policy paradigm for climate change adaptation. *Environ. Policy Gov.*, **19**, 358–373, doi:10.1002/eet.523.
- Collins, M., and Coauthors, 2013: Long-term Climate Change: Projections, Commitments and Irreversibility. *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, G.-K.P. T. F. Stocker, D. Qin, Ed., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- Conant, R. T., Cerri, C. E. P., Osborne, B. B. & Paustian, K. Grassland management impacts on soil carbon stocks: A new synthesis: A. *Ecol. Appl.*(2017).
- Conant, R. T., and K. Paustian, 2002: Potential soil carbon sequestration in overgrazed grassland ecosystems. *Global Biogeochem. Cycles*, **16**, 90-1-90–99, doi:10.1029/2001GB001661. <http://doi.wiley.com/10.1029/2001GB001661> (Accessed November 12, 2018).
- , ———, S. J. Del Grosso, and W. J. Parton, 2005: Nitrogen pools and fluxes in grassland soils sequestering carbon. *Nutr. Cycl. Agroecosystems*, **71**, 239–248, doi:10.1007/s10705-004-5085-z. <http://link.springer.com/10.1007/s10705-004-5085-z> (Accessed April 12, 2019).
- , C. E. P. Cerri, B. B. Osborne, and K. Paustian, 2017: Grassland management impacts on soil carbon stocks: a new synthesis. *Ecol. Appl.*, **27**, 662–668, doi:10.1002/eap.1473. <http://doi.wiley.com/10.1002/eap.1473> (Accessed June 5, 2018).
- de Coninck, H., and Coauthors, 2018: Strengthening and implementing the global response. *Global Warming of 1.5C: an IPCC special report on the impacts of global warming of 1.5C above pre-industrial levels and related global greenhouse gas emission pathways, in the context of strengthening the global response to the threat of climate change* <http://www.ipcc.ch/report/sr15/>.
- Cools, J., D. Innocenti, and S. O’Brien, 2016: Lessons from flood early warning systems. *Environ. Sci.*

- Policy*, **58**, 117–122, doi:10.1016/J.ENVSCI.2016.01.006.
<https://www.sciencedirect.com/science/article/pii/S1462901116300065> (Accessed June 1, 2018).
- Coomes, O. T., and Coauthors, 2015a: Farmer seed networks make a limited contribution to agriculture? Four common misconceptions. *Food Policy*, **56**, 41–50, doi:10.1016/J.FOODPOL.2015.07.008.
<https://www.sciencedirect.com/science/article/pii/S030691921500086X> (Accessed June 1, 2018).
- , and Coauthors, 2015b: Farmer seed networks make a limited contribution to agriculture? Four common misconceptions. *Food Policy*, **56**, 41–50, doi:10.1016/J.FOODPOL.2015.07.008.
<https://www.sciencedirect.com/science/article/pii/S030691921500086X> (Accessed November 11, 2018).
- Corbera, E., C. Hunsberger, and C. Vaddhanaphuti, 2017: Climate change policies, land grabbing and conflict: perspectives from Southeast Asia. *Can. J. Dev. Stud. Can. D ETUDES DU DEVELOPEMENT*, **38**, 297–304, doi:10.1080/02255189.2017.1343413.
- Correa, D. F., H. L. Beyer, H. P. Possingham, S. R. Thomas-Hall, and P. M. Schenk, 2017: Biodiversity impacts of bioenergy production: Microalgae vs. first generation biofuels. *Renew. Sustain. Energy Rev.*, **74**, 1131–1146, doi:10.1016/J.RSER.2017.02.068.
<https://www.sciencedirect.com/science/article/pii/S1364032117302691> (Accessed April 5, 2019).
- Cossalter, C., and C. Pye-Smith, 2003: *Fast-wood forestry: myths and realities*.
<https://books.google.com/books?hl=en&lr=&id=fu3uciRDD2UC&oi=fnd&pg=PR4&dq=Cossalter+C.,+Pye-Smith+C.+Fast-Wood+Forestry:+Myths+and+Realities.+Center+for+International+Forestry+Research,+Jakarta,+2003.+60+&ots=ZIDhBWj5Hu&sig=EkcdWvuh0zjHsC6PVgWYLid8VYc> (Accessed November 11, 2018).
- Costanza, R., and Coauthors, 1997: The value of the world’s ecosystem services and natural capital. *Nature*, **387**, 253–260, doi:10.1038/387253a0. <http://www.nature.com/articles/387253a0> (Accessed April 4, 2019).
- Coutts, A., and R. Harris, 2013: *A multi-scale assessment of urban heating in Melbourne during an extreme heat event and policy approaches for adaptation*. 64 pp.
- Couwenberg, J., R. Dommain, and H. Joosten, 2010: Greenhouse gas fluxes from tropical peatlands in south-east Asia. *Glob. Chang. Biol.*, **16**, 1715–1732, doi:10.1111/j.1365-2486.2009.02016.x.
- Coveney, J., and L. A. O’Dwyer, 2009: Effects of mobility and location on food access. *Health Place*, **15**, 45–55, doi:10.1016/J.HEALTHPLACE.2008.01.010.
<https://www.sciencedirect.com/science/article/abs/pii/S1353829208000178> (Accessed April 15, 2019).
- CRED, 2015: *The Human Cost of Natural Disasters 2015*. Geneva, Switzerland,.
- Creutzig, F., and Coauthors, 2015: Bioenergy and climate change mitigation: An assessment. *GCB Bioenergy*, **7**, 916–944, doi:10.1111/gcbb.12205.
- Cromsigt, J. P. G. M., M. te Beest, G. I. H. Kerley, M. Landman, E. le Roux, and F. A. Smith, 2018: Trophic rewilding as a climate change mitigation strategy? *Philos. Trans. R. Soc. B Biol. Sci.*, **373**, 20170440, doi:10.1098/rstb.2017.0440.
<http://www.royalsocietypublishing.org/doi/10.1098/rstb.2017.0440> (Accessed April 13, 2019).
- Crush, J. S., and G. B. Frayne, 2011: Urban food insecurity and the new international food security agenda. *Dev. South. Afr.*, **28**, 527–544, doi:10.1080/0376835X.2011.605571.
<http://www.tandfonline.com/doi/abs/10.1080/0376835X.2011.605571> (Accessed April 15, 2019).
- Cuéllar, A. D., and M. E. Webber, 2010: Wasted Food, Wasted Energy: The Embedded Energy in Food Waste in the United States. *Environ. Sci. Technol.*, **44**, 6464–6469, doi:10.1021/es100310d.

- <https://pubs.acs.org/doi/10.1021/es100310d> (Accessed April 16, 2019).
- Curtis, P. G., C. M. Slay, N. L. Harris, A. Tyukavina, and M. C. Hansen, 2018: Classifying drivers of global forest loss. *Science* (80-.), **361**, 1108–1111, doi:10.1126/science.aau3445. <http://www.sciencemag.org/lookup/doi/10.1126/science.aau3445> (Accessed November 11, 2018).
- D’Amato, A. W., J. B. Bradford, S. Fraver, and B. J. Palik, 2011: Forest management for mitigation and adaptation to climate change: Insights from long-term silviculture experiments. *For. Ecol. Manage.*, **262**, 803–816, doi:10.1016/j.foreco.2011.05.014.
- D’Odorico, P., Y. He, S. Collins, S. F. J. De Wekker, V. Engel, and J. D. Fuentes, 2013: Vegetation-microclimate feedbacks in woodland-grassland ecotones. *Glob. Ecol. Biogeogr.*, **22**, 364–379, doi:10.1111/geb.12000. <http://doi.wiley.com/10.1111/geb.12000> (Accessed June 5, 2018).
- Dagar, J., P. Sharma, D. Sharma, and A. Singh, 2016a: *Innovative saline agriculture*. <https://link.springer.com/content/pdf/10.1007/978-81-322-2770-0.pdf> (Accessed November 8, 2018).
- Dagar, J. C., D. K. Sharma, P. C. Sharma, and A. K. Singh, 2016b: *Innovative saline agriculture*. 1-519 pp.
- Dai, Z., 2010: Intensive agropastoralism: dryland degradation, the Grain-to-Green Program and islands of sustainability in the Mu Us Sandy Land of China. *Agric. Ecosyst. Environ.*, **138**, 249–256. <https://www.sciencedirect.com/science/article/pii/S0167880910001453> (Accessed November 11, 2018).
- Daioglou, V., J. C. Doelman, B. Wicke, A. Faaij, and D. P. van Vuuren, 2019: Integrated assessment of biomass supply and demand in climate change mitigation scenarios. *Glob. Environ. Chang.*, **54**, 88–101, doi:10.1016/J.GLOENVCHA.2018.11.012. <https://www.sciencedirect.com/science/article/pii/S0959378018303765> (Accessed April 14, 2019).
- Dale, V. H., and Coauthors, 2001: Climate Change and Forest Disturbances Climate change can affect forests by altering the frequency, intensity, duration, and timing of fire, drought, introduced species, insect and pathogen outbreaks, hurricanes, windstorms, ice storms, or landslides. *Bioscience*, **51**, 723–734, doi:10.1641/0006-3568(2001)051[0723:ccafd]2.0.co;2. <https://academic.oup.com/bioscience/article/51/9/723/288247> (Accessed April 4, 2019).
- Dallimer, M., L. C. Stringer, S. E. Orchard, P. Osano, G. Njoroge, C. Wen, and P. Gicheru, 2018: Who uses sustainable land management practices and what are the costs and benefits? Insights from Kenya. *L. Degrad. Dev.*, **29**, 2822–2835, doi:10.1002/ldr.3001. <http://doi.wiley.com/10.1002/ldr.3001> (Accessed November 9, 2018).
- Danley, B., and C. Widmark, 2016: Evaluating conceptual definitions of ecosystem services and their implications. *Ecol. Econ.*, **126**, 132–138, doi:10.1016/J.ECOLECON.2016.04.003. <https://www.sciencedirect.com/science/article/pii/S0921800915300549> (Accessed April 4, 2019).
- DARA, 2012: *Climate Vulnerability Monitor*. <https://daraint.org/climate-vulnerability-monitor/climate-vulnerability-monitor-2012/report/>.
- Darnton-Hill, I., and B. Cogill, 2010: Maternal and Young Child Nutrition Adversely Affected by External Shocks Such As Increasing Global Food Prices. *J. Nutr.*, **140**, 162S–169S, doi:10.3945/jn.109.111682. <https://academic.oup.com/jn/article/140/1/162S/4600327> (Accessed November 11, 2018).
- Dasgupta, P., J. F. Morton, D. Dodman, B. Karapinar, F. Meza, M. G. Rivera-Ferre, A. Toure Sarr, and K. E. Vincent, 2014: Rural Areas. *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, p. 151.

- Datta, K. K., C. De Jong, and O. P. Singh, 2000: Reclaiming salt-affected land through drainage in Haryana, India: A financial analysis. *Agric. Water Manag.*, **46**, 55–71, doi:10.1016/S0378-3774(00)00077-9.
- Davies, J., 2016: Enabling Governance for Sustainable Land Management. *Land Restoration*, Academic Press, 67–76 <https://www.sciencedirect.com/science/article/pii/B9780128012314000069> (Accessed November 10, 2018).
- Davis, S. C., and Coauthors, 2013a: Management swing potential for bioenergy crops. *GCB Bioenergy*, doi:10.1111/gcbb.12042.
- , and Coauthors, 2013b: Management swing potential for bioenergy crops. *GCB Bioenergy*, doi:10.1111/gcbb.12042.
- Deal, B., and D. Schunk, 2004: Spatial dynamic modeling and urban land use transformation: a simulation approach to assessing the costs of urban sprawl. *Ecol. Econ.*, **51**, 79–95, doi:10.1016/j.ecolecon.2004.04.008. <http://linkinghub.elsevier.com/retrieve/pii/S0921800904002241> (Accessed November 11, 2018).
- DeCicco, J. M., 2013: Biofuel’s carbon balance: doubts, certainties and implications. *Clim. Change*, 801–814.
- Delgado, C. L., 2010: Sources of growth in smallholder agriculture integration of smallholders with processors in sub-saharan Africa: the role of vertical and marketers of high value-added items. doi:10.1080/03031853.1999.9524913. <https://www.tandfonline.com/action/journalInformation?journalCode=ragr20> (Accessed April 15, 2019).
- Demaiily, K.-E., and S. Darly, 2017: Urban agriculture on the move in Paris: The routes of temporary gardening in the neoliberal city. *ACME An Int. E-Journal Crit. Geogr.*, <https://hal.archives-ouvertes.fr/hal-01972331/> (Accessed April 16, 2019).
- Deng, F. F., and Y. Huang, 2004: Uneven land reform and urban sprawl in China: the case of Beijing. *Prog. Plann.*, **61**, 211–236, doi:10.1016/j.progress.2003.10.004. www.elsevier.com/locate/plplann (Accessed April 15, 2019).
- DERM, 2011: Salinity management handbook Second edition. 188.
- Derpsch, R., T. Friedrich, A. Kassam, and H. Li, 2010: Current Status of Adoption of No-till Farming in the World and Some of its Main Benefits. *Int. J. Agric. Biol. Eng.*, **3**, 1–25, doi:10.25165/IJABE.V3I1.223. <https://ijabe.org/index.php/ijabe/article/view/223> (Accessed May 31, 2018).
- Descheemaeker, K., S. J. Oosting, S. Homann-Kee Tui, P. Masikati, G. N. Falconnier, and K. E. Giller, 2016: Climate change adaptation and mitigation in smallholder crop–livestock systems in sub-Saharan Africa: a call for integrated impact assessments. *Reg. Environ. Chang.*, **16**, 2331–2343, doi:10.1007/s10113-016-0957-8. <http://link.springer.com/10.1007/s10113-016-0957-8> (Accessed June 1, 2018).
- Dhingra, C., S. Gandhi, A. Chaurey, and P. K. Agarwal, 2008: Access to clean energy services for the urban and peri-urban poor: a case-study of Delhi, India. *Energy Sustain. Dev.*, **12**, 49–55, doi:10.1016/S0973-0826(09)60007-7. <https://www.sciencedirect.com/science/article/pii/S0973082609600077> (Accessed April 15, 2019).
- Díaz, S., and Coauthors, 2015: The IPBES Conceptual Framework — connecting nature and people. *Curr. Opin. Environ. Sustain.*, **14**, 1–16, doi:10.1016/J.COSUST.2014.11.002. <https://www.sciencedirect.com/science/article/pii/S187734351400116X> (Accessed April 4, 2019).

- , and Coauthors, 2018: Assessing nature's contributions to people. *Science (80-.)*, **359**, 270–272, doi:10.1126/science.aap8826.
- Dickie, I. A., and Coauthors, 2014: Conflicting values: ecosystem services and invasive tree management. *Biol. Invasions*, **16**, 705–719, doi:10.1007/s10530-013-0609-6. <http://link.springer.com/10.1007/s10530-013-0609-6> (Accessed April 12, 2019).
- Dickinson, D., L. Balduccio, J. Buysse, F. Ronsse, G. van Huylenbroeck, and W. Prins, 2014: Cost-benefit analysis of using biochar to improve cereals agriculture. *GCB Bioenergy*, **7**, 850–864, doi:10.1111/gcbb.12180. <http://doi.wiley.com/10.1111/gcbb.12180> (Accessed August 13, 2018).
- Dietz, T., E. Ostrom, and P. Stern, 2013: The struggle to govern the commons. *Science (80-.)*, **302**, 1907–1912. <http://science.sciencemag.org/content/302/5652/1907.short> (Accessed November 10, 2018).
- DiGiano, M. L., and A. E. Racelis, 2012: Robustness, adaptation and innovation: Forest communities in the wake of Hurricane Dean. *Appl. Geogr.*, **33**, 151–158, doi:10.1016/j.apgeog.2011.10.004.
- Dillon, P., and M. Arshad, 2016: Managed aquifer recharge in integrated water resource management. *Integrated Groundwater Management*, A.J. Jakeman, O. Barreteau, R.J. Hunt, J.-D. Rinaudo, and A. Ross, Eds., Springer, Cham, 435–452 <http://link.springer.com/10.1007/978-3-319-23576-9>.
- Dixon, J., A. M. Omwega, S. Friel, C. Burns, K. Donati, and R. Carlisle, 2007: The Health Equity Dimensions of Urban Food Systems. *J. Urban Heal.*, **84**, 118–129, doi:10.1007/s11524-007-9176-4. <http://link.springer.com/10.1007/s11524-007-9176-4> (Accessed April 15, 2019).
- Djalante, R., F. Thomalla, M. S. Sinapoy, and M. Carnegie, 2012: Building resilience to natural hazards in Indonesia: Progress and challenges in implementing the Hyogo Framework for Action. *Nat. Hazards*, **62**, 779–803, doi:10.1007/s11069-012-0106-8.
- Djeddaoui, F., M. Chadli, R. G.-R. Sensing, and undefined 2017, Desertification Susceptibility Mapping Using Logistic Regression Analysis in the Djelfa Area, Algeria. *mdpi.com.*, <https://www.mdpi.com/2072-4292/9/10/1031> (Accessed November 12, 2018).
- Djenontin, I., S. Foli, and L. Zulu, 2018: Revisiting the Factors Shaping Outcomes for Forest and Landscape Restoration in Sub-Saharan Africa: A Way Forward for Policy, Practice and Research. *Sustainability*, **10**, 906. <https://www.mdpi.com/2071-1050/10/4/906> (Accessed November 10, 2018).
- Djordjević, S., D. Butler, P. Gourbesville, O. Mark, and E. Pasche, 2011: New policies to deal with climate change and other drivers impacting on resilience to flooding in urban areas: the CORFU approach. *Environ. Sci. Policy*, **14**, 864–873, doi:10.1016/J.ENVSCL.2011.05.008. <https://www.sciencedirect.com/science/article/pii/S1462901111000748> (Accessed April 15, 2019).
- Doelman, J. C., and Coauthors, 2018: Exploring SSP land-use dynamics using the IMAGE model: Regional and gridded scenarios of land-use change and land-based climate change mitigation. *Glob. Environ. Chang.*, **48**, 119–135, doi:10.1016/J.GLOENVCHA.2017.11.014. <https://www.sciencedirect.com/science/article/pii/S0959378016306392> (Accessed April 5, 2019).
- Doerr, S., and C. Santín, 2016: Global trends in wildfire and its impacts: perceptions versus realities in a changing world. *Philos. Trans. R. Soc. B Biol. Sci.*, **371**, 20150345. <http://rstb.royalsocietypublishing.org/content/371/1696/20150345.abstract> (Accessed November 12, 2018).
- Dolan, C., and K. Sorby, 2003: *Gender and employment in high-value agriculture industries*.
- Don, A., and Coauthors, 2012: Land-use change to bioenergy production in Europe: implications for the greenhouse gas balance and soil carbon. *Glob. Chang. Biol. Bioenergy*, **4**, 372–391, doi:10.1111/j.1757-1707.2011.01116.x.

- Doney, S. C., V. J. Fabry, R. A. Feely, and J. A. Kleypas, 2009: Ocean Acidification: The Other CO₂ Problem. *Ann. Rev. Mar. Sci.*, **1**, 169–192, doi:10.1146/annurev.marine.010908.163834. <http://www.annualreviews.org/doi/10.1146/annurev.marine.010908.163834> (Accessed April 15, 2019).
- Dooley, K., and S. Kartha, 2018a: Land-based negative emissions: risks for climate mitigation and impacts on sustainable development. *Int. Environ. Agreements Polit. Law Econ.*, **18**, 79–98, doi:10.1007/s10784-017-9382-9. <http://link.springer.com/10.1007/s10784-017-9382-9> (Accessed June 2, 2018).
- Dornburg, V., and Coauthors, 2010: Bioenergy revisited: Key factors in global potentials of bioenergy. *Energy Environ. Sci.*, **3**, 258–267, doi:10.1039/B922422J. <http://dx.doi.org/10.1039/B922422J>.
- Dorward, L. J., 2012: Where are the best opportunities for reducing greenhouse gas emissions in the food system (including the food chain)? A comment. *Food Policy*, doi:10.1016/j.foodpol.2012.04.006.
- Dresner, M., C. Handelman, S. Braun, and G. Rollwagen-Bollens, 2015: Environmental identity, pro-environmental behaviors, and civic engagement of volunteer stewards in Portland area parks. *Environ. Educ. Res.*, **21**, 991–1010, doi:10.1080/13504622.2014.964188.
- Drewry, J., 2006: Natural recovery of soil physical properties from treading damage of pastoral soils in New Zealand and Australia: a review. *Agric. Ecosyst. Environ.*, **114**, 159–169. <https://www.sciencedirect.com/science/article/pii/S0167880905005621> (Accessed November 11, 2018).
- Du, B., K. Zhang, G. Song, and Z. Wen, 2006: Methodology for an urban ecological footprint to evaluate sustainable development in China. *Int. J. Sustain. Dev. World Ecol.*, **13**, 245–254, doi:10.1080/13504500609469676. <https://www.tandfonline.com/doi/full/10.1080/13504500609469676> (Accessed April 15, 2019).
- Duarte, C. M., I. J. Losada, I. E. Hendriks, I. Mazarrasa, and N. Marbà, 2013: The role of coastal plant communities for climate change mitigation and adaptation. *Nat. Clim. Chang.*, **3**, 961–968, doi:10.1038/nclimate1970. <http://www.nature.com/articles/nclimate1970> (Accessed April 4, 2019).
- Dudley, N., S. Stolton, A. Belokurov, L. Krueger, N. Lopoukhine, K. MacKinnon, T. Sandwith, and N. Sekhran, 2010: *Natural Solutions: Protected areas helping people cope with climate change*. Gland, Switzerland, Washington DC and New York, USA, 130 pp.
- Dugan, A. J., R. Birdsey, V. S. Mascorro, M. Magnan, C. E. Smyth, M. Olguin, and W. A. Kurz, 2018: A systems approach to assess climate change mitigation options in landscapes of the United States forest sector. *Carbon Balance Manag.*, **13**, 13, doi:10.1186/s13021-018-0100-x. <https://cbmjournal.biomedcentral.com/articles/10.1186/s13021-018-0100-x> (Accessed April 5, 2019).
- Durigan, G., N. Guerin, and J. N. M. N. da Costa, 2013: Ecological restoration of Xingu Basin headwaters: motivations, engagement, challenges and perspectives. *Philos. Trans. R. Soc. B Biol. Sci.*, **368**, 20120165–20120165, doi:10.1098/rstb.2012.0165.
- Ebi, K. L., and J. K. Schmier, 2005: A Stitch in Time: Improving Public Health Early Warning Systems for Extreme Weather Events. *Epidemiol. Rev.*, **27**, 115–121, doi:10.1093/epirev/mxi006. <http://academic.oup.com/epirev/article/27/1/115/520819/A-Stitch-in-Time-Improving-Public-Health-Early> (Accessed April 15, 2019).
- Edelenbosch, O. Y., and Coauthors, 2017: Decomposing passenger transport futures: Comparing results of global integrated assessment models. *Transp. Res. Part D Transp. Environ.*, **55**, 281–293, doi:10.1016/j.trd.2016.07.003. <https://linkinghub.elsevier.com/retrieve/pii/S1361920916301304> (Accessed November 11, 2018).

- Edenhofer, O., and Coauthors, 2011: *Renewable energy sources and climate change mitigation: Special report of the intergovernmental panel on climate change*.
- Edmonds, J., and Coauthors, 2013a: Can radiative forcing be limited to 2.6 Wm⁻² without negative emissions from bioenergy AND CO₂ capture and storage? *Clim. Change*, **118**, doi:10.1007/s10584-012-0678-z.
- Edmonds, J., and Coauthors, 2013b: Can radiative forcing be limited to 2.6 Wm⁻² without negative emissions from bioenergy AND CO₂ capture and storage? *Clim. Change*, **118**, 29–43, doi:10.1007/s10584-012-0678-z. <http://link.springer.com/10.1007/s10584-012-0678-z> (Accessed November 11, 2018).
- Ellis, E. C., and N. Ramankutty, 2008: Putting people in the map: anthropogenic biomes of the world. *Front. Ecol. Environ.*, **6**, 439–447, doi:10.1890/070062. <http://doi.wiley.com/10.1890/070062> (Accessed June 2, 2018).
- Ellis, F., 1998: Household strategies and rural livelihood diversification. *J. Dev. Stud.*, **35**, 1–38, doi:10.1080/00220389808422553. <http://www.tandfonline.com/doi/abs/10.1080/00220389808422553> (Accessed November 11, 2018).
- , 2008: The Determinants of Rural Livelihood Diversification in Developing Countries. *J. Agric. Econ.*, **51**, 289–302, doi:10.1111/j.1477-9552.2000.tb01229.x. <http://doi.wiley.com/10.1111/j.1477-9552.2000.tb01229.x> (Accessed November 11, 2018).
- , and J. Sumberg, 1998: Food production, urban areas and policy responses. *World Dev.*, **26**, 213–225, doi:10.1016/S0305-750X(97)10042-0. <https://www.sciencedirect.com/science/article/abs/pii/S0305750X97100420> (Accessed April 15, 2019).
- Ellison, D., and Coauthors, 2017a: Trees, forests and water: Cool insights for a hot world. *Glob. Environ. Chang.*, **43**, 51–61, doi:10.1016/j.gloenvcha.2017.01.002.
- .
- Elmqvist, B., and L. Olsson, 2007: Livelihood diversification: continuity and change in the Sahel. *GeoJournal*, **67**, 167–180, doi:10.1007/s10708-007-9043-6. <http://link.springer.com/10.1007/s10708-007-9043-6> (Accessed April 15, 2019).
- Emerton, L. (Lucy), J. Bishop, L. Thomas, Germany. Bundesamt für Naturschutz., IUCN World Commission on Protected Areas., James Cook University., and Cooperative Research Centre for Tropical Rainforest Ecology and Management., 2006: *Sustainable financing of protected areas: a global review of challenges and options*. IUCN--the World Conservation Union, 97 pp. <https://books.google.co.nz/books?hl=en&lr=&id=BQIOWj8iE3EC&oi=fnd&pg=PP11&dq=Sustainable+Financing+of+Protected+Areas:+A+Global+Review+of+Challenges+and+Options.+&ots=kuh4AzzKeP&sig=lgIhARr4uTYldwO-zrOy4P6OjVw#v=onepage&q=Sustainable+Financing+of+Protected+Areas%3A+A+Global+Review+of+Challenges+and+Options.&f=false> (Accessed April 4, 2019).
- Enarson, E., and L. Meyreles, 2004: International perspectives on gender and disaster: differences and possibilities. *Int. J. Sociol. Soc. Policy*, **24**, 49–93, doi:10.1108/01443330410791064. <https://www.emeraldinsight.com/doi/10.1108/01443330410791064> (Accessed April 15, 2019).
- Ensor, J., R. Berger, and S. Huq, 2014: *Community-based adaptation to climate change: Emerging lessons*. Practical Action Publishing, Rugby, 183--197 pp.
- Epron, D., C. Plain, F.-K. Ndiaye, P. Bonnaud, C. Pasquier, and J. Ranger, 2016: Effects of compaction by heavy machine traffic on soil fluxes of methane and carbon dioxide in a temperate broadleaved forest. *For. Ecol. Manage.*, **382**, 1–9, doi:10.1016/J.FORECO.2016.09.037.

- <https://www.sciencedirect.com/science/article/pii/S037811271630648X> (Accessed May 31, 2018).
- Erb, K. H., and Coauthors, 2018: Unexpectedly large impact of forest management and grazing on global vegetation biomass. *Nature*, **553**, 73–76, doi:10.1038/nature25138.
- Erb, K.-H., H. Haberl, and C. Plutzer, 2012: Dependency of global primary bioenergy crop potentials in 2050 on food systems, yields, biodiversity conservation and political stability. *Energy Policy*, **47**, 260–269, doi:https://doi.org/10.1016/j.enpol.2012.04.066. <http://www.sciencedirect.com/science/article/pii/S0301421512003710>.
- , C. Lauk, T. Kastner, A. Mayer, M. C. Theurl, and H. Haberl, 2016: Exploring the biophysical option space for feeding the world without deforestation. *Nat. Commun.*, **7**, 11382. <http://dx.doi.org/10.1038/ncomms11382>.
- , and Coauthors, 2017: Unexpectedly large impact of forest management and grazing on global vegetation biomass. *Nature*, **553**, 73–76, doi:10.1038/nature25138. <http://www.nature.com/doi/10.1038/nature25138> (Accessed November 11, 2018).
- Ercilla-Montserrat, M., D. Sanjuan-Delmás, E. Sanyé-Mengual, L. Calvet-Mir, K. Banderas, J. Rieradevall, and X. Gabarrell, 2019: Analysis of the consumer's perception of urban food products from a soilless system in rooftop greenhouses: a case study from the Mediterranean area of Barcelona (Spain). *Agric. Human Values*, 1–19, doi:10.1007/s10460-019-09920-7. <http://link.springer.com/10.1007/s10460-019-09920-7> (Accessed April 12, 2019).
- Eriksson, L. O., and Coauthors, 2012: Climate change mitigation through increased wood use in the European construction sector—towards an integrated modelling framework. *Eur. J. For. Res.*, **131**, 131–144, doi:10.1007/s10342-010-0463-3. <https://doi.org/10.1007/s10342-010-0463-3>.
- Erismann, J. W., M. A. Sutton, J. Galloway, Z. Klimont, and W. Winiwarter, 2008: How a century of ammonia synthesis changed the world. *Nat. Geosci.*, **1**, 636–639, doi:10.1038/ngeo325. <http://www.nature.com/articles/ngeo325> (Accessed April 17, 2019).
- Erwin, K. L., 2009: Wetlands and global climate change: the role of wetland restoration in a changing world. *Wetl. Ecol. Manag.*, **17**, 71–84, doi:10.1007/s11273-008-9119-1. <http://link.springer.com/10.1007/s11273-008-9119-1> (Accessed June 5, 2018).
- Esteves, T., and Coauthors, 2012: Mitigating land degradation caused by wildfire: application of the PESERA model to fire-affected sites in central Portugal. *Geoderma*, **191**, 40–50. <https://www.sciencedirect.com/science/article/pii/S001670611200016X> (Accessed November 11, 2018).
- Estudillo, J. P., and K. Otsuka, 1999: Green Revolution, Human Capital, and Off-Farm Employment: Changing Sources of Income among Farm Households in Central Luzon, 1966–1994. *Econ. Dev. Cult. Change*, **47**, 497–523, doi:10.1086/452417. <https://www.journals.uchicago.edu/doi/10.1086/452417> (Accessed April 15, 2019).
- Evans, R. G., and E. J. Sadler, 2008: Methods and technologies to improve efficiency of water use. *Water Resour. Res.*, **44**, 1–15, doi:10.1029/2007WR006200. <http://doi.wiley.com/10.1029/2007WR006200> (Accessed June 5, 2018).
- Evers, M., and S. Hofmeister, 2011: Gender mainstreaming and participative planning for sustainable land management. *J. Environ. Plan. Manag.*, **54**, 1315–1329, doi:10.1080/09640568.2011.573978. <http://www.tandfonline.com/doi/abs/10.1080/09640568.2011.573978> (Accessed November 10, 2018).
- Fabinyi, M., L. Evans, and S. J. Foale, 2014: Social-ecological systems, social diversity, and power: insights from anthropology and political ecology. *Ecol. Soc.*, **19**, art28, doi:10.5751/ES-07029-190428. <http://www.ecologyandsociety.org/vol19/iss4/art28/> (Accessed November 10, 2018).

- Fafchamps, M., C. Udry, and K. Czukas, 1998: Drought and saving in West Africa: are livestock a buffer stock? *J. Dev. Econ.*, **55**, 273–305, doi:10.1016/S0304-3878(98)00037-6. <https://www.sciencedirect.com/science/article/pii/S0304387898000376> (Accessed June 1, 2018).
- Fakhruddin, S. H. M., A. Kawasaki, and M. S. Babel, 2015: Community responses to flood early warning system: Case study in Kaijuri Union, Bangladesh. *Int. J. Disaster Risk Reduct.*, **14**, 323–331, doi:10.1016/J.IJDRR.2015.08.004. <https://www.sciencedirect.com/science/article/pii/S2212420915300509> (Accessed June 1, 2018).
- Fan, S., R. Pandya-Lorch, and International Food Policy Research Institute., 2012: *Reshaping agriculture for nutrition and health*. International Food Policy Research Institute, 213 pp. <https://books.google.co.nz/books?hl=en&lr=&id=gVQZXr38z5sC&oi=fnd&pg=PP1&dq=Reshaping+agriculture+for+nutrition+and+health&ots=0De5YX8puh&sig=UsyQazTMyVrxIwahp2uRNt0tN5I#v=onepage&q=Reshaping+agriculture+for+nutrition+and+health&f=false> (Accessed April 16, 2019).
- FAO, 2006: *Fire management: voluntary guidelines. Principles and strategic actions*. Rome, Italy,.
- , 2011: *The State of the World's land and water resources for Food and Agriculture. Managing systems at risk*. London, UK,.
- , 2015: *The impact of disasters on agriculture and food security*. ITALY,.
- , 2017a: *The future of food and agriculture - Trends and challenges*. Rome, 180 pp.
- , 2017b: *Voluntary Guidelines for Sustainable Soil Management*. <http://www.fao.org/3/a-bl813e.pdf> (Accessed November 7, 2018).
- , and Coauthors, 2015: *World ' s Soil Resources*. 650 pp.
- FAO and ITPS, 2015: *Status of the World's Soil Resources (SWSR) - Main Report*.
- Favero, A., and E. Massetti, 2014: Trade of woody biomass for electricity generation under climate mitigation policy. *Resour. Energy Econ.*, **36**, 166–190, doi:10.1016/J.RESENEECO.2013.11.005. <https://www.sciencedirect.com/science/article/pii/S0928765513000808> (Accessed April 14, 2019).
- , and R. Mendelsohn, 2014: Using Markets for Woody Biomass Energy to Sequester Carbon in Forests. *J. Assoc. Environ. Resour. Econ.*, **1**, 75–95, doi:10.1086/676033. <http://www.journals.uchicago.edu/doi/10.1086/676033> (Accessed December 20, 2017).
- , and —, 2017: The Land-Use Consequences of Woody Biomass with More Stringent Climate Mitigation Scenarios. *J. Environ. Prot. (Irvine, Calif.)*, **8**, 61–73, doi:10.4236/jep.2017.81006. <http://www.scirp.org/journal/doi.aspx?DOI=10.4236/jep.2017.81006>.
- Fawcett, R., B. Christensen, and D. Tierney, 1994: The impact of conservation tillage on pesticide runoff into surface water: a review and analysis. *J. Soil Water Conserv.*, **49**, 126–135.
- Feng, X., B. Fu, S. Piao, S. Wang, P. Ciais, ... Z. Z.-N. C., and undefined 2016, Revegetation in China's Loess Plateau is approaching sustainable water resource limits. *nature.com.*, <https://www.nature.com/articles/nclimate3092> (Accessed November 12, 2018).
- Fenton, A., J. Paavola, and A. Tallontire, 2017: The Role of Microfinance in Household Livelihood Adaptation in Satkhira District, Southwest Bangladesh. *World Dev.*, **92**, 192–202, doi:10.1016/j.worlddev.2016.12.004.
- Fiksel, J., 2006: A framework for sustainable materials management. *JOM*, **58**, 15–22, doi:10.1007/s11837-006-0047-3. <http://link.springer.com/10.1007/s11837-006-0047-3> (Accessed April 15, 2019).
- Findell, K. L., A. Berg, P. Gentine, J. P. Krasting, B. R. Lintner, S. Malyshev, J. A. Santanello, and E.

- Shevliakova, 2017: The impact of anthropogenic land use and land cover change on regional climate extremes. *Nat. Commun.*, **8**, 989, doi:10.1038/s41467-017-01038-w.
- Flannigan, M., B. Stocks, M. Turetsky, and M. Wotton, 2009: Impacts of climate change on fire activity and fire management in the circumboreal forest. *Glob. Chang. Biol.*, **15**, 549–560, doi:10.1111/j.1365-2486.2008.01660.x. <http://doi.wiley.com/10.1111/j.1365-2486.2008.01660.x> (Accessed April 4, 2019).
- Fletschner, D., and L. Kenney, 2014: Rural Women’s Access to Financial Services: Credit, Savings, and Insurance. *Gender in Agriculture*, Springer, Netherlands, Dordrecht, 187–208.
- Foley, J., N. Ramankutty, K. Brauman, E. C.- Nature, and undefined 2011, Solutions for a cultivated planet. *nature.com*,. <https://www.nature.com/articles/nature10452> (Accessed November 8, 2018).
- Foley, J. A., and Coauthors, 2005: Global Consequences of Land Use. *Science (80-.)*, **309**, 570–574, doi:10.1126/SCIENCE.1111772. <http://www.ncbi.nlm.nih.gov/pubmed/16040698> (Accessed April 5, 2019).
- , and Coauthors, 2011: Solutions for a cultivated planet. *Nature*, **478**, 337–342, doi:10.1038/nature10452. <http://www.nature.com/articles/nature10452> (Accessed November 11, 2018).
- ForestEurope, 2016: *Sustainable Forest Management Implementation*. <https://foresteurope.org/sustainable-forest-management-implementation/>.
- Forsyth, T., 2013: Community-based adaptation: A review of past and future challenges. *Wiley Interdiscip. Rev. Clim. Chang.*, **4**, 439–446, doi:10.1002/wcc.231.
- Fothergill, A., and L. A. Peek, 2004: Poverty and Disasters in the United States: A Review of Recent Sociological Findings. *Nat. Hazards*, **32**, 89–110, doi:10.1023/B:NHAZ.0000026792.76181.d9. <http://link.springer.com/10.1023/B:NHAZ.0000026792.76181.d9> (Accessed April 15, 2019).
- Francis, C. A., T. E. Hansen, A. A. Fox, P. J. Hesje, H. E. Nelson, A. E. Lawseth, and A. English, 2012a: Farmland conversion to non-agricultural uses in the US and Canada: current impacts and concerns for the future. *Int. J. Agric. Sustain.*, **10**, 8–24, doi:10.1080/14735903.2012.649588. <https://www.tandfonline.com/doi/full/10.1080/14735903.2012.649588> (Accessed June 4, 2018).
- , ———, ———, ———, ———, ———, and ———, 2012b: Farmland conversion to non-agricultural uses in the US and Canada: current impacts and concerns for the future. *Int. J. Agric. Sustain.*, **10**, 8–24, doi:10.1080/14735903.2012.649588. <http://www.tandfonline.com/doi/abs/10.1080/14735903.2012.649588> (Accessed June 4, 2018).
- Frank, S., E. Schmid, P. Havlik, U. A. Schneider, H. Böttcher, J. Balkovič, and M. Obersteiner, 2015: The dynamic soil organic carbon mitigation potential of European cropland. *Glob. Environ. Chang.*, **35**, 269–278, doi:https://doi.org/10.1016/j.gloenvcha.2015.08.004. <http://www.sciencedirect.com/science/article/pii/S095937801530025X>.
- , and Coauthors, 2017: Reducing greenhouse gas emissions in agriculture without compromising food security? *Environ. Res. Lett.*, **12**, 105004, doi:10.1088/1748-9326/aa8c83. <http://stacks.iop.org/1748-9326/12/i=10/a=105004?key=crossref.000b6ab1748d1b0af07f66e0f496f0a3>.
- Frankl, A., J. Poesen, M. Haile, J. Deckers, and J. Nyssen, 2013: Quantifying long-term changes in gully networks and volumes in dryland environments: The case of Northern Ethiopia. *Geomorphology*, **201**, 254–263, doi:10.1016/J.GEOMORPH.2013.06.025. <https://www.sciencedirect.com/science/article/pii/S0169555X13003474> (Accessed June 2, 2018).
- Franks, J., 2010: Boundary organizations for sustainable land management: The example of Dutch

- Environmental Co-operatives. *Ecol. Econ.*, **70**, 283–295, doi:10.1016/j.ecolecon.2010.08.011. <http://linkinghub.elsevier.com/retrieve/pii/S0921800910003356> (Accessed November 10, 2018).
- Franz, M., N. Schlitz, and K. P. Schumacher, 2017: Globalization and the water-energy-food nexus - Using the global production networks approach to analyze society-environment relations. *Environmental Science and Policy*.
- Freeman, J., L. Kobziar, E. W. Rose, and W. Cropper, 2017a: A critique of the historical-fire-regime concept in conservation. *Conserv. Biol.*, **31**, 976–985, doi:10.1111/cobi.12942. <http://doi.wiley.com/10.1111/cobi.12942> (Accessed May 31, 2018).
- , —, —, and —, 2017b: A critique of the historical-fire-regime concept in conservation. *Conserv. Biol.*, **31**, 976–985, doi:10.1111/cobi.12942. <http://doi.wiley.com/10.1111/cobi.12942> (Accessed April 18, 2019).
- Freudenberg, N., S. Galea, and D. Vlahov, 2005: Beyond Urban Penalty and Urban Sprawl: Back to Living Conditions as the Focus of Urban Health. *J. Community Health*, **30**, 1–11, doi:10.1007/s10900-004-6091-4. <http://link.springer.com/10.1007/s10900-004-6091-4> (Accessed April 15, 2019).
- Fricko, O., and Coauthors, 2017: The marker quantification of the Shared Socioeconomic Pathway 2: A middle-of-the-road scenario for the 21st century. *Glob. Environ. Chang.*, **42**, 251–267, doi:https://doi.org/10.1016/j.gloenvcha.2016.06.004. <http://www.sciencedirect.com/science/article/pii/S0959378016300784>.
- Fridahl, M., and M. Lehtveer, 2018: Bioenergy with carbon capture and storage (BECCS): Global potential, investment preferences, and deployment barriers. doi:10.1016/j.erss.2018.03.019.
- Friel, S., M. Marmot, A. J. McMichael, T. Kjellstrom, and D. Vågerö, 2008: Global health equity and climate stabilisation: a common agenda. *Lancet*, **372**, 1677–1683, doi:10.1016/S0140-6736(08)61692-X. <https://www.sciencedirect.com/science/article/pii/S014067360861692X> (Accessed April 15, 2019).
- Frumkin, H., 2002: Urban Sprawl and Public Health. *Public Health Rep.*, **117**, 201–217, doi:10.1093/phr/117.3.201. <http://phr.oupjournals.org/cgi/doi/10.1093/phr/117.3.201> (Accessed November 11, 2018).
- Fujimori, S., T. Hasegawa, T. Masui, ... K. T.-G. E., and undefined 2017, SSP3: AIM implementation of shared socioeconomic pathways. *Elsevier*.
- Fujimori, S., T. Hasegawa, T. Masui, K. Takahashi, D. S. Herran, H. Dai, Y. Hijioka, and M. Kainuma, 2017a: SSP3: AIM implementation of Shared Socioeconomic Pathways. *Glob. Environ. Chang.*, **42**, 268–283, doi:10.1016/j.gloenvcha.2016.06.009. <https://linkinghub.elsevier.com/retrieve/pii/S0959378016300838> (Accessed November 11, 2018).
- , —, —, —, —, —, —, and —, 2017b: SSP3: AIM implementation of Shared Socioeconomic Pathways. *Glob. Environ. Chang.*, **42**, 268–283, doi:https://doi.org/10.1016/j.gloenvcha.2016.06.009. <http://www.sciencedirect.com/science/article/pii/S0959378016300838>.
- , and Coauthors, 2018a: A multi-model assessment of food security implications of well below 2°C scenarios. *Nat. Commun.*, **In Review**.
- Fujimori, S., T. Hasegawa, J. Rogelj, ... X. S.-E., and U. 2018, 2018b: Inclusive climate change mitigation and food security policy under 1.5° C climate goal. *Environ. Res. Lett.*, **13**. <http://iopscience.iop.org/article/10.1088/1748-9326/aad0f7/meta> (Accessed November 11, 2018).
- Fuss, S., W. Lamb, ... M. C.-E., and undefined 2018, Negative emissions—Part 2: Costs, potentials and

- side effects. *iopscience.iop.org*. <http://iopscience.iop.org/article/10.1088/1748-9326/aabf9f/meta> (Accessed November 8, 2018).
- Fuss, S., and Coauthors, 2014: Betting on negative emissions. *Nat. Clim. Chang.*, **4**, 850, doi:10.1038/nclimate2392.
- , and Coauthors, 2018b: Negative emissions - Part 2: Costs, potentials and side effects. *Environ. Res. Lett.*, **13**, 2–4, doi:<https://doi.org/10.1088/1748-9326/aabf9f>.
- Gabriel, K. M. A., and W. R. Endlicher, 2011: Urban and rural mortality rates during heat waves in Berlin and Brandenburg, Germany. *Environ. Pollut.*, **159**, 2044–2050, doi:10.1016/J.ENVPOL.2011.01.016. <https://www.sciencedirect.com/science/article/pii/S0269749111000388> (Accessed June 2, 2018).
- Galal, O., M. Corroon, and C. Tirado, 2010: Urban Environment and Health: Food Security. *Asia Pacific J. Public Heal.*, **22**, 254S–261S, doi:10.1177/1010539510372993. <http://journals.sagepub.com/doi/10.1177/1010539510372993> (Accessed April 16, 2019).
- Gao, B., and Coauthors, 2018: Chinese cropping systems are a net source of greenhouse gases despite soil carbon sequestration. *Glob. Chang. Biol.*, doi:10.1111/gcb.14425. <http://doi.wiley.com/10.1111/gcb.14425> (Accessed November 10, 2018).
- Gao, X.; Li, H.; Zhao, X.; Ma, W. & Wu, P. 2018. [Identifying a suitable revegetation technique for soil restoration on water-limited and degraded land: Considering both deep soil moisture deficit and soil organic carbon sequestration](#) - *Geoderma*, 2319, 61-69
- Gao, L., and B. A. Bryan, 2017: Finding pathways to national-scale land-sector sustainability. *Nature*, **544**, 217–222, doi:10.1038/nature21694. <http://www.nature.com/doi/10.1038/nature21694> (Accessed November 10, 2018).
- Garbrecht, J., M. Nearing, J. Steiner, ... X. Z.-W. and C., and undefined 2015, Can conservation trump impacts of climate change on soil erosion? An assessment from winter wheat cropland in the Southern Great Plains of the United. *Elsevier*.
- , ——, ——, X. Zhang, and M. Nichols, 2015: Can conservation trump impacts of climate change on soil erosion? An assessment from winter wheat cropland in the Southern Great Plains of the United. *Weather Clim. Extrem.*, **10**, 32–39. <https://www.sciencedirect.com/science/article/pii/S2212094715300037> (Accessed November 11, 2018).
- Garcia, A. S., and T. Wanner, 2017: Gender inequality and food security: lessons from the gender-responsive work of the International Food Policy Research Institute and the Bill and Melinda Gates Foundation. *Food Secur.*, **9**, 1091–1103, doi:10.1007/s12571-017-0718-7. <http://link.springer.com/10.1007/s12571-017-0718-7> (Accessed November 10, 2018).
- Garcia, D. J., and F. You, 2016: The water-energy-food nexus and process systems engineering: A new focus. *Comput. Chem. Eng.*, **91**, 49–67, doi:10.1016/J.COMPCHEMENG.2016.03.003. <https://www.sciencedirect.com/science/article/pii/S0098135416300552> (Accessed April 16, 2019).
- Gardiner, M. M., and Coauthors, 2009: Landscape diversity enhances biological control of an introduced crop pest in the north-central USA. *Ecol. Appl.*, **19**, 143–154, doi:10.1890/07-1265.1. <http://doi.wiley.com/10.1890/07-1265.1> (Accessed April 16, 2019).
- Gariano, S., and F. Guzzetti, 2016a: Landslides in a changing climate. *Earth-Science Rev.*, **162**, 227–252. <https://www.sciencedirect.com/science/article/pii/S0012825216302458> (Accessed November 11, 2018).

- Gariano, S. L., and F. Guzzetti, 2016b: Landslides in a changing climate. *Earth-Science Rev.*, **162**, 227–252, doi:10.1016/J.EARSCIREV.2016.08.011. <https://www.sciencedirect.com/science/article/pii/S0012825216302458> (Accessed April 4, 2019).
- Garnett, S., and Coauthors, 2018: A spatial overview of the global importance of Indigenous lands for conservation. *Nat. Sustain.*, **1**, 369.
- Garnett, T., 2011: Where are the best opportunities for reducing greenhouse gas emissions in the food system (including the food chain)? *Food Policy*, **36**, S23–S32. <https://www.sciencedirect.com/science/article/pii/S0306919210001132> (Accessed May 31, 2018).
- Garnett, T., and Coauthors, 2013a: Sustainable Intensification in Agriculture: Premises and Policies. *Science (80-.)*, **341**.
- , and Coauthors, 2013b: Sustainable Intensification in Agriculture: Premises and Policies. *Science*, **341**, 33–34, doi:10.1126/science.1234485. <http://www.ncbi.nlm.nih.gov/pubmed/21885781> (Accessed August 10, 2018).
- Garschagen, M., 2016: Decentralizing urban disaster risk management in a centralized system? Agendas, actors and contentions in Vietnam. *Habitat Int.*, **52**, 43–49, doi:10.1016/j.habitatint.2015.08.030.
- Gassert, F., M. Luck, M. Landis, P. Reig, T. S.-W. R. Institute, and U. 2014, 2014: Aqueduct global maps 2.1: Constructing decision-relevant global water risk indicators. *wri.org.*, https://www.wri.org/sites/default/files/Aqueduct_Global_Maps_2.1-Constructing_Decision-Relevant_Global_Water_Risk_Indicators_final_0.pdf (Accessed November 10, 2018).
- Gaveau, D. L. A., D. Sheil, Husnayaen, M. A. Salim, S. Arjasakusuma, M. Ancrenaz, P. Pacheco, and E. Meijaard, 2016: Rapid conversions and avoided deforestation: examining four decades of industrial plantation expansion in Borneo. *Sci. Rep.*, **6**, 32017, doi:10.1038/srep32017. <http://www.nature.com/articles/srep32017> (Accessed June 2, 2018).
- Gebremeskel, G., T. G. Gebremicael, and A. Girmay, 2018: Economic and environmental rehabilitation through soil and water conservation, the case of Tigray in northern Ethiopia. *J. Arid Environ.*, **151**, 113–124, doi:10.1016/J.JARIDENV.2017.12.002. <https://www.sciencedirect.com/science/article/pii/S014019631730232X> (Accessed June 2, 2018).
- Genesio, L., and Coauthors, 2011: Early warning systems for food security in West Africa: evolution, achievements and challenges. *Atmos. Sci. Lett.*, **12**, 142–148, doi:10.1002/asl.332. <http://doi.wiley.com/10.1002/asl.332> (Accessed November 11, 2018).
- Gengenbach, H., R. A. Schurman, T. J. Bassett, W. A. Munro, and W. G. Moseley, 2018: Limits of the New Green Revolution for Africa: Reconceptualising gendered agricultural value chains. *Geogr. J.*, **184**, 208–214, doi:10.1111/geoj.12233. <http://doi.wiley.com/10.1111/geoj.12233> (Accessed April 15, 2019).
- Gerlak, A. K., and T. Heikkila, 2011: Building a Theory of Learning in Collaboratives: Evidence from the Everglades Restoration Program. *J. Public Adm. Res. Theory*, **21**, 619–644, doi:10.1093/jopart/muq089. <https://academic.oup.com/jpart/article-lookup/doi/10.1093/jopart/muq089> (Accessed November 10, 2018).
- Gibbs, H. K., and J. M. Salmon, 2015: Mapping the world's degraded lands. *Appl. Geogr.*, **57**, 12–21, doi:10.1016/J.APGEOG.2014.11.024. <https://www.sciencedirect.com/science/article/pii/S0143622814002793> (Accessed June 6, 2018).
- Gibson, J., G. Boe-Gibson, and G. Stichbury, 2015: Urban land expansion in India 1992–2012. *Food Policy*, **56**, 100–113, doi:10.1016/J.FOODPOL.2015.08.002. <https://www.sciencedirect.com/science/article/pii/S0306919215001013> (Accessed May 31, 2018).

- Gilbert, C. L., 2012a: International agreements to manage food price volatility. *Glob. Food Sec.*, **1**, 134–142, doi:10.1016/J.GFS.2012.10.001. <https://www.sciencedirect.com/science/article/pii/S2211912412000168> (Accessed June 1, 2018).
- , 2012b: International agreements to manage food price volatility. *Glob. Food Sec.*, **1**, 134–142, doi:10.1016/j.gfs.2012.10.001. <https://linkinghub.elsevier.com/retrieve/pii/S2211912412000168> (Accessed April 18, 2019).
- Gilbert, C. L., and C. W. Morgan, 2010: Food price volatility. *Philos. Trans. R. Soc. B Biol. Sci.*, **365**, 3023–3034, doi:10.1098/rstb.2010.0139. <http://www.royalsocietypublishing.org/doi/10.1098/rstb.2010.0139> (Accessed April 15, 2019).
- Gill, J. C., and B. D. Malamud, 2017a: Anthropogenic processes, natural hazards, and interactions in a multi-hazard framework. *Earth-Science Rev.*, **166**, 246–269, doi:10.1016/J.EARSCIREV.2017.01.002. <https://www.sciencedirect.com/science/article/pii/S0012825216302227> (Accessed June 5, 2018).
- , and ———, 2017b: Anthropogenic processes, natural hazards, and interactions in a multi-hazard framework. *Earth-Science Rev.*, **166**, 246–269, doi:10.1016/j.earscirev.2017.01.002. <https://linkinghub.elsevier.com/retrieve/pii/S0012825216302227> (Accessed April 18, 2019).
- Gimona, A., L. Poggio, I. Brown, and M. Castellazzi, 2012: Woodland networks in a changing climate: Threats from land use change. *Biol. Conserv.*, **149**, 93–102, doi:10.1016/J.BIOCON.2012.01.060. <https://www.sciencedirect.com/science/article/pii/S0006320712000730> (Accessed April 4, 2019).
- Giné, X., R. Townsend, and J. Vickery, 2008: *Patterns of rainfall insurance participation in rural India*. The World Bank, 47 pp.
- Gioli, G., T. Khan, S. Bisht, and J. Scheffran, 2014: Migration as an Adaptation Strategy and its Gendered Implications: A Case Study From the Upper Indus Basin. *Mt. Res. Dev.*, **34**, 255–265, doi:10.1659/MRD-JOURNAL-D-13-00089.1. <http://www.bioone.org/doi/10.1659/MRD-JOURNAL-D-13-00089.1> (Accessed April 15, 2019).
- Glauber, J. W., 2004: Crop Insurance Reconsidered. *Am. J. Agric. Econ.*, **86**, 1179–1195, doi:10.1111/j.0002-9092.2004.00663.x. <https://academic.oup.com/ajae/article-lookup/doi/10.1111/j.0002-9092.2004.00663.x> (Accessed April 15, 2019).
- Global Panel on Agriculture and Food Systems for Nutrition, 2016: *Food systems and diets: Facing the challenges of the 21st century*. London, UK, 133 pp.
- Göbel, C., N. Langen, A. Blumenthal, P. Teitscheid, and G. Ritter, 2015: Cutting food waste through cooperation along the food supply chain. *Sustainability*, **7**, 1429–1445. <http://www.mdpi.com/2071-1050/7/2/1429/htm> (Accessed May 1, 2018).
- Godfray, H. C. J., and T. Garnett, 2014: Food security and sustainable intensification. *Philos. Trans. R. Soc. B Biol. Sci.*, **369**, 20120273–20120273, doi:10.1098/rstb.2012.0273. <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2012.0273> (Accessed April 15, 2019).
- , and Coauthors, 2010a: Food Security: The Challenge of Feeding 9 Billion People. *Science (80-.)*, **327**, 812–818, doi:10.1126/science.1185383. <http://www.sciencemag.org/cgi/doi/10.1126/science.1185383> (Accessed November 11, 2018).
- Godfray, H. C. J., and Coauthors, 2010b: Food security: The challenge of feeding 9 billion people. *Science (80-.)*, 1185383, doi:10.1126/science.1185383.
- Godschalk, D. R., 2003: Urban Hazard Mitigation: Creating Resilient Cities. *Nat. Hazards Rev.*, **4**, 136–143, doi:10.1061/(ASCE)1527-6988(2003)4:3(136). <http://ascelibrary.org/doi/10.1061/%28ASCE%291527-6988%282003%294%3A3%28136%29>

(Accessed April 15, 2019).

- Goldstein, B., M. Hauschild, J. Fernandez, and M. Birkved, 2016: Testing the environmental performance of urban agriculture as a food supply in northern climates. *J. Clean. Prod.*, **135**, 984–994. <https://www.sciencedirect.com/science/article/pii/S0959652616308952> (Accessed November 11, 2018).
- Gomiero, T., M. G. Paoletti, and D. Pimentel, 2008: Energy and Environmental Issues in Organic and Conventional Agriculture. *CRC. Crit. Rev. Plant Sci.*, **27**, 239–254, doi:10.1080/07352680802225456. <https://www.tandfonline.com/doi/full/10.1080/07352680802225456> (Accessed April 15, 2019).
- Goodwin, B. K., and V. H. Smith, 2003: An Ex Post Evaluation of the Conservation Reserve, Federal Crop Insurance, and Other Government Programs: Program Participation and Soil Erosion. *J. Agric. Resour. Econ.*, **28**, 201–216, doi:10.2307/40987182. <http://www.jstor.org/stable/40987182> (Accessed June 1, 2018).
- Goodwin, B. K., and V. H. Smith, 2013: What Harm Is Done By Subsidizing Crop Insurance? *Am. J. Agric. Econ.*, **95**, 489–497, doi:10.1093/ajae/aas092. <https://academic.oup.com/ajae/article-lookup/doi/10.1093/ajae/aas092> (Accessed June 1, 2018).
- , M. L. Vandever, and J. L. Deal, 2004: An Empirical Analysis of Acreage Effects of Participation in the Federal Crop Insurance Program. *Am. J. Agric. Econ.*, **86**, 1058–1077, doi:10.1111/j.0002-9092.2004.00653.x. <https://academic.oup.com/ajae/article-lookup/doi/10.1111/j.0002-9092.2004.00653.x> (Accessed June 1, 2018).
- Gorham, E., 1991: Northern Peatlands: Role in the Carbon Cycle and Probable Responses to Climatic Warming. *Ecol. Appl.*, **1**, 182–195, doi:10.2307/1941811. <http://doi.wiley.com/10.2307/1941811> (Accessed April 4, 2019).
- Gough, N. E. V. and C., 2016: Expert assessment concludes negative emissions scenarios may not deliver. *Environ. Res. Lett.*, **11**, 95003. <http://stacks.iop.org/1748-9326/11/i=9/a=095003>.
- Gould, K. A., 2014: Everyday expertise: land regularization and the conditions for landgrabs in Peten, Guatemala. *Environ. Plan. A*, **46**, 2353–2368, doi:10.1068/a140188p.
- Graham-Rowe, E., D. C. Jessop, and P. Sparks, 2014: Identifying motivations and barriers to minimising household food waste. *Resour. Conserv. Recycl.*, **84**, 15–23, doi:https://doi.org/10.1016/j.resconrec.2013.12.005. <http://www.sciencedirect.com/science/article/pii/S0921344913002711>.
- Graham, N. T., and Coauthors, 2018a: Water Sector Assumptions for the Shared Socioeconomic Pathways in an Integrated Modeling framework. *Water Resour. Res.*, **0**, doi:10.1029/2018WR023452. <https://doi.org/10.1029/2018WR023452>.
- Graham, N. T., and Coauthors, 2018b: Water Sector Assumptions for the Shared Socioeconomic Pathways in an Integrated Modeling Framework. *Water Resour. Res.*, **54**, 6423–6440, doi:10.1029/2018WR023452. <http://doi.wiley.com/10.1029/2018WR023452> (Accessed November 10, 2018).
- Graham, V., S. G. Laurance, A. Grech, and O. Venter, 2017: Spatially explicit estimates of forest carbon emissions, mitigation costs and REDD+ opportunities in Indonesia. *Environ. Res. Lett.*, **12**, 44017, doi:10.1088/1748-9326/aa6656. <http://stacks.iop.org/1748-9326/12/i=4/a=044017?key=crossref.969b6ca7c50dd7aa01dc8119235e7322> (Accessed June 2, 2018).
- Grassi, G., J. House, F. Dentener, S. Federici, M. den Elzen, and J. Penman, 2017: The key role of forests in meeting climate targets requires science for credible mitigation. *Nat. Clim. Chang.*, **7**, 220,

doi:10.1038/nclimate3227.

- , R. Pilli, J. House, S. Federici, and W. A. Kurz, 2018: Science-based approach for credible accounting of mitigation in managed forests. *Carbon Balance Manag.*, **13**, 8, doi:10.1186/s13021-018-0096-2.
- Greene, R., W. Timms, P. Rengasamy, M. Arshad, and R. Cresswell, 2016: Soil and Aquifer Salinization: Toward an Integrated Approach for Salinity Management of Groundwater. *Integrated Groundwater Management*, A.J. Jakeman, O. Barreteau, R.J. Hunt, J.-D. Rinaudo, and A. Ross, Eds., Springer, 377–412 <http://link.springer.com/10.1007/978-3-319-23576-9>.
- Greenough, G., M. McGeehin, S. M. Bernard, J. Trtanj, J. Riad, and D. Engelberg, 2001: The potential impacts of climate variability and change on health impacts of extreme weather events in the United States. *Environ. Health Perspect.*, **109**, 191–198, doi:10.1289/ehp.109-1240666. <https://ehp.niehs.nih.gov/doi/10.1289/ehp.109-1240666> (Accessed April 15, 2019).
- Greenwood, K., and B. McKenzie, 2001: Grazing effects on soil physical properties and the consequences for pastures: a review. *Aust. J. Exp. Agric.*, **41**, 1231–1250. <http://www.publish.csiro.au/AN/EA00102> (Accessed November 11, 2018).
- Grey, D., and C. W. Sadoff, 2007: Sink or Swim? Water security for growth and development. *Water Policy*, **9**, 545–571, doi:10.2166/wp.2007.021.
- Grey, S., and R. Patel, 2015: Food sovereignty as decolonization: some contributions from Indigenous movements to food system and development politics. *Agric. Human Values*, **32**, 431–444, doi:10.1007/s10460-014-9548-9. <http://link.springer.com/10.1007/s10460-014-9548-9> (Accessed April 15, 2019).
- Grimm, N. B., S. H. Faeth, N. E. Golubiewski, C. L. Redman, J. Wu, X. Bai, and J. M. Briggs, 2008: Global Change and the Ecology of Cities. *Science (80-.)*, **319**, 756–760, doi:10.1126/SCIENCE.1150195. https://science.sciencemag.org/content/319/5864/756?casa_token=_KKfbn0uetIAAAAA:Nuc9zz_R3n9hBuuAv6JXsHUephJwM0LEx57Sj32BEQKuGOVnkD042aep3RZ6ZLkt-tKCjqmripuc67yw (Accessed April 15, 2019).
- Griscom, B. W., and Coauthors, 2017: Natural climate solutions. *Proc. Natl. Acad. Sci.*, **114**, 11645–11650, doi:10.1073/pnas.1710465114. <http://www.pnas.org/lookup/doi/10.1073/pnas.1710465114> (Accessed November 8, 2018).
- de Groot, R. S., M. A. Wilson, and R. M. . Boumans, 2002: A typology for the classification, description and valuation of ecosystem functions, goods and services. *Ecol. Econ.*, **41**, 393–408, doi:10.1016/S0921-8009(02)00089-7. <https://www.sciencedirect.com/science/article/pii/S0921800902000897> (Accessed April 4, 2019).
- Gross, C. L., M. Fatemi, and I. H. Simpson, 2017: Seed provenance for changing climates: early growth traits of nonlocal seed are better adapted to future climatic scenarios, but not to current field conditions. *Restor. Ecol.*, **25**, 577–586, doi:10.1111/rec.12474. <http://doi.wiley.com/10.1111/rec.12474> (Accessed April 4, 2019).
- Del Grosso, S., P. Smith, M. Galdos, A. Hastings, and W. Parton, 2014: Sustainable energy crop production. *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2014.07.007.
- Grubler, A., and Coauthors, 2018: A low energy demand scenario for meeting the 1.5 °C target and sustainable development goals without negative emission technologies. *Nat. Energy*, **3**, 515–527, doi:10.1038/s41560-018-0172-6. <http://dx.doi.org/10.1038/s41560-018-0172-6>.
- Guan, D., and K. Hubacek, 2007: Assessment of regional trade and virtual water flows in China. *Ecol. Econ.*, **61**, 159–170, doi:10.1016/J.ECOLECON.2006.02.022.

- <https://www.sciencedirect.com/science/article/pii/S0921800906000930> (Accessed April 16, 2019).
- Gunatilake, H., D. Roland-Holst, and G. Sugiyarto, 2014: Energy security for India: Biofuels, energy efficiency and food productivity. *Energy Policy*, **65**, 761–767. <https://www.sciencedirect.com/science/article/pii/S0301421513010732> (Accessed November 11, 2018).
- Guo, X., H. Dong, C. Yang, Q. Zhang, C. Liao, F. Zha, and L. Gao, 2016: Application of goethite modified biochar for tylosin removal from aqueous solution. *Colloids Surfaces A Physicochem. Eng. Asp.*, **502**, 81–88, doi:10.1016/J.COLSURFA.2016.05.015. <https://www.sciencedirect.com/science/article/pii/S0927775716303235> (Accessed June 6, 2018).
- Gurwick, N. P., L. A. Moore, C. Kelly, and P. Elias, 2013: A Systematic Review of Biochar Research, with a Focus on Its Stability in situ and Its Promise as a Climate Mitigation Strategy. *PLoS One*, **8**, e75932, doi:10.1371/journal.pone.0075932. <http://dx.plos.org/10.1371/journal.pone.0075932> (Accessed April 4, 2019).
- Gustavsson, J., C. Cederberg, U. Sonesson, R. van Otterdijk, and A. Meybeck, 2011: *Global food losses and food waste - Extent, causes and prevention*. Rome, Italy,.
- Gustavsson, L., and R. Sathre, 2011: Energy and CO2 analysis of wood substitution in construction. *Clim. Change*, **105**, 129–153, doi:10.1007/s10584-010-9876-8. <http://link.springer.com/10.1007/s10584-010-9876-8> (Accessed April 15, 2019).
- Gustavsson, L., and Coauthors, 2006: *THE ROLE OF WOOD MATERIAL FOR GREENHOUSE GAS MITIGATION*. 1097-1127 pp.
- Haberl, H., K. H. Erb, F. Krausmann, A. Bondeau, C. Lauk, C. Müller, C. Plutzer, and J. K. Steinberger, 2011: Global bioenergy potentials from agricultural land in 2050: Sensitivity to climate change, diets and yields. *Biomass and Bioenergy*, doi:10.1016/j.biombioe.2011.04.035.
- Hadley, C., D. Lindstrom, F. Tessema, and T. Belachew, 2008: Gender bias in the food insecurity experience of Ethiopian adolescents. *Soc. Sci. Med.*, **66**, 427–438, doi:10.1016/J.SOCSCIMED.2007.08.025. <https://www.sciencedirect.com/science/article/abs/pii/S0277953607004856> (Accessed April 15, 2019).
- de Haen, H., and V. Réquillart, 2014: Linkages between sustainable consumption and sustainable production: some suggestions for foresight work. *Food Secur.*, **6**, 87–100, doi:10.1007/s12571-013-0323-3. <http://link.springer.com/10.1007/s12571-013-0323-3> (Accessed April 15, 2019).
- Haggblade, S., P. Hazell, and T. Reardon, 2010: The Rural Non-farm Economy: Prospects for Growth and Poverty Reduction. *World Dev.*, **38**, 1429–1441, doi:10.1016/J.WORLDDEV.2009.06.008. <https://www.sciencedirect.com/science/article/abs/pii/S0305750X10000963> (Accessed April 15, 2019).
- , N. M. Me-Nsope, and J. M. Staatz, 2017: Food security implications of staple food substitution in Sahelian West Africa. *Food Policy*, **71**, 27–38, doi:10.1016/J.FOODPOL.2017.06.003. <https://www.sciencedirect.com/science/article/pii/S0306919216302366> (Accessed June 1, 2018).
- Hallegatte, S., 2012: A Cost Effective Solution to Reduce Disaster Losses in Developing Countries: Hydro-Meteorological Services, Early Warning, and Evacuation. https://papers.ssrn.com/sol3/papers.cfm?abstract_id=2051341 (Accessed June 1, 2018).
- Hallegatte, S., and Coauthors, 2016: *Shock Waves: Managing the Impacts of Climate Change on Poverty*. The World Bank,.
- Halloran, A., J. Clement, N. Kornum, C. Bucatariu, and J. Magid, 2014: Addressing food waste reduction

- in Denmark. *Food Policy*, **49**, 294–301, doi:10.1016/J.FOODPOL.2014.09.005. <https://www.sciencedirect.com/science/article/abs/pii/S0306919214001365> (Accessed April 15, 2019).
- Hammill, A., R. Matthew, and E. McCarter, 2008: Microfinance and climate change adaptation. *IDS Bull.*, **39**, 113–122.
- Hamza, M. A., and W. K. Anderson, 2005a: Soil compaction in cropping systems: A review of the nature, causes and possible solutions. *Soil Tillage Res.*, **82**, 121–145, doi:10.1016/J.STILL.2004.08.009. <https://www.sciencedirect.com/science/article/pii/S0167198704001849> (Accessed May 31, 2018).
- , and ———, 2005b: Soil compaction in cropping systems: A review of the nature, causes and possible solutions. *Soil Tillage Res.*, **82**, 121–145, doi:10.1016/J.STILL.2004.08.009. <https://www.sciencedirect.com/science/article/pii/S0167198704001849> (Accessed November 9, 2018).
- Hanasaki, N., and Coauthors, 2013a: A global water scarcity assessment under Shared Socio-economic Pathways - Part 2: Water availability and scarcity. *Hydrol. Earth Syst. Sci.*, **17**, 2393–2413, doi:10.5194/hess-17-2393-2013.
- , and Coauthors, 2013b: A global water scarcity assessment under Shared Socio-economic Pathways &—; Part 2: Water availability and scarcity. *Hydrol. Earth Syst. Sci.*, **17**, 2393–2413, doi:10.5194/hess-17-2393-2013. <https://www.hydrol-earth-syst-sci.net/17/2393/2013/> (Accessed November 10, 2018).
- Hanjra, M. A., and M. E. Qureshi, 2010: Global water crisis and future food security in an era of climate change. *Food Policy*, **35**, 365–377, doi:10.1016/J.FOODPOL.2010.05.006. <https://www.sciencedirect.com/science/article/abs/pii/S030691921000059X> (Accessed April 16, 2019).
- Hansen, M., and Coauthors, 2013a: High-Resolution Global Maps of 21st-Century Forest Cover Change. *Science (80-.)*, **342**, 850–853, doi:10.1126/science.1070656. <http://www.ncbi.nlm.nih.gov/pubmed/12169731> (Accessed June 2, 2018).
- Hansen, M. C., and Coauthors, 2013b: High-Resolution Global Maps of 21st-Century Forest Cover Change. *Science (80-.)*, **342**, 850–853, doi:10.1126/science.1244693. <http://www.sciencemag.org/cgi/doi/10.1126/science.1244693> (Accessed November 11, 2018).
- Harada, Y., T. H. Whitlow, J. Russell-Anelli, M. T. Walter, N. L. Bassuk, and M. A. Rutzke, 2019: The heavy metal budget of an urban rooftop farm. *Sci. Total Environ.*, **660**, 115–125, doi:10.1016/J.SCITOTENV.2018.12.463. <https://www.sciencedirect.com/science/article/pii/S0048969718353592> (Accessed April 12, 2019).
- Haregeweyn, N., and Coauthors, 2015: Soil erosion and conservation in Ethiopia. *Prog. Phys. Geogr.*, **39**, 750–774, doi:10.1177/0309133315598725. <http://journals.sagepub.com/doi/10.1177/0309133315598725> (Accessed November 11, 2018).
- Harper, A. B., and Coauthors, 2018: Land-use emissions play a critical role in land-based mitigation for Paris climate targets. *Nat. Commun.*, **9**, 2938, doi:10.1038/s41467-018-05340-z. <https://doi.org/10.1038/s41467-018-05340-z>.
- Harris, E., T. Ladreiter-Knauss, K. Butterbach-Bahl, B. Wolf, and M. Bahn, 2018: Land-use and abandonment alters methane and nitrous oxide fluxes in mountain grasslands. *Sci. Total Environ.*, doi:10.1016/j.scitotenv.2018.02.119.
- Harris, Z. M., R. Spake, and G. Taylor, 2015: Land use change to bioenergy: A meta-analysis of soil carbon and GHG emissions. *Biomass and Bioenergy*, doi:10.1016/j.biombioe.2015.05.008.

- Harrison, P. A., R. W. Dunford, I. P. Holman, and M. D. A. Rounsevell, 2016: Climate change impact modelling needs to include cross-sectoral interactions. *Nat. Clim. Chang.*, doi:10.1038/nclimate3039.
- Hasan, J., S. States, And, and R. Deininger, 2009: Safeguarding The Security Of Public Water Supplies Using Early Warning Systems: A Brief Review. *J. Contemp. Water Res. Educ.*, **129**, 27–33, doi:10.1111/j.1936-704X.2004.mp129001007.x. <http://doi.wiley.com/10.1111/j.1936-704X.2004.mp129001007.x> (Accessed April 16, 2019).
- Hasegawa, T., S. Fujimori, Y. Shin, A. Tanaka, K. Takahashi, and T. Masui, 2015a: Consequence of Climate Mitigation on the Risk of Hunger. *Environ. Sci. Technol.*, **49**, 7245–7253, doi:10.1021/es5051748. <http://pubs.acs.org/doi/10.1021/es5051748> (Accessed November 10, 2018).
- , ———, ———, ———, ———, and ———, 2015b: Consequence of Climate Mitigation on the Risk of Hunger. *Environ. Sci. Technol.*, doi:10.1021/es5051748.
- , and Coauthors, 2018: Risk of increased food insecurity under stringent global climate change mitigation policy. *Nat. Clim. Chang.*, **8**, 699–703, doi:10.1038/s41558-018-0230-x. <https://doi.org/10.1038/s41558-018-0230-x>.
- Havemenn, T., and V. Muccione, 2011: *Mechanisms for agricultural climate change mitigation incentives for smallholders*. CCAFS Report no. 6. Copenhagen, Denmark,.
- Havlik, P., and Coauthors, 2014: Climate change mitigation through livestock system transitions. *Proc. Natl. Acad. Sci.*, **111**, 3709–3714, doi:10.1073/pnas.1308044111.
- Hawken, P. *Project Drawdown: The most comprehensive plan ever proposed to reverse global warming*. (Penguin Books, 2017).
- He, B., Y. Cai, W. Ran, X. Zhao, and H. Jiang, 2015: Spatiotemporal heterogeneity of soil salinity after the establishment of vegetation on a coastal saline field. *Catena*, **127**, 129–134, doi:10.1016/j.catena.2014.12.028. <http://dx.doi.org/10.1016/j.catena.2014.12.028>.
- Headey, D., and S. Fan, 2008: Anatomy of a crisis: the causes and consequences of surging food prices. *Agric. Econ.*, **39**, 375–391, doi:10.1111/j.1574-0862.2008.00345.x. <http://doi.wiley.com/10.1111/j.1574-0862.2008.00345.x> (Accessed November 11, 2018).
- Hearn, M. D., T. Baranowski, J. Baranowski, C. Doyle, M. Smith, L. S. Lin, and K. Resnicow, 1998a: Environmental Influences on Dietary Behavior among Children: Availability and Accessibility of Fruits and Vegetables Enable Consumption. *J. Heal. Educ.*, **29**, 26–32, doi:10.1080/10556699.1998.10603294. <https://www.tandfonline.com/doi/full/10.1080/10556699.1998.10603294> (Accessed May 30, 2018).
- , ———, ———, ———, ———, ———, and ———, 1998b: Environmental Influences on Dietary Behavior among Children: Availability and Accessibility of Fruits and Vegetables Enable Consumption. *J. Heal. Educ.*, **29**, 26–32, doi:10.1080/10556699.1998.10603294. <https://www.tandfonline.com/doi/full/10.1080/10556699.1998.10603294> (Accessed April 18, 2019).
- Hebrok, M., and C. Boks, 2017: Household food waste: Drivers and potential intervention points for design – An extensive review. *J. Clean. Prod.*, **151**, 380–392, doi:10.1016/J.JCLEPRO.2017.03.069. <https://www.sciencedirect.com/science/article/pii/S0959652617305048> (Accessed April 15, 2019).
- Heck, V., D. Gerten, W. Lucht, and A. Popp, 2018: Biomass-based negative emissions difficult to reconcile with planetary boundaries. *Nat. Clim. Chang.*, **8**, 151–155, doi:10.1038/s41558-017-0064-y. <http://www.nature.com/articles/s41558-017-0064-y> (Accessed June 2, 2018).
- Hedenus, F., Wirsenius, S. & Johansson, D. J. A. The importance of reduced meat and dairy consumption for meeting stringent climate change targets. *Clim.*

Change 124, 79–91 (2014).

- Heikkila, T., and A. K. Gerlak, 2018: Working on learning: how the institutional rules of environmental governance matter. *J. Environ. Plan. Manag.*, 1–18, doi:10.1080/09640568.2018.1473244. <https://www.tandfonline.com/doi/full/10.1080/09640568.2018.1473244> (Accessed November 10, 2018).
- Hejazi, M., and Coauthors, 2014a: Long-term global water projections using six socioeconomic scenarios in an integrated assessment modeling framework. *Technol. Forecast. Soc. Change*, **81**, 205–226, doi:10.1016/j.techfore.2013.05.006. <http://dx.doi.org/10.1016/j.techfore.2013.05.006>.
- , and Coauthors, 2014b: Long-term global water projections using six socioeconomic scenarios in an integrated assessment modeling framework. *Technol. Forecast. Soc. Change*, **81**, 205–226, doi:10.1016/j.techfore.2013.05.006. <http://dx.doi.org/10.1016/j.techfore.2013.05.006> (Accessed November 10, 2018).
- Hejazi, M. I., and Coauthors, 2014c: Integrated assessment of global water scarcity over the 21st century under multiple climate change mitigation policies. *Hydrol. Earth Syst. Sci.*, **18**, 2859–2883, doi:10.5194/hess-18-2859-2014.
- Hejazi, M. I., and Coauthors, 2015a: 21st century United States emissions mitigation could increase water stress more than the climate change it is mitigating. *Proc. Natl. Acad. Sci.*, **112**, 10635–10640, doi:10.1073/pnas.1421675112. <http://www.pnas.org/lookup/doi/10.1073/pnas.1421675112> (Accessed November 11, 2018).
- , and Coauthors, 2015b: 21st century United States emissions mitigation could increase water stress more than the climate change it is mitigating. *Proc. Natl. Acad. Sci.*, **112**, 10635–10640, doi:10.1073/pnas.1421675112. <http://www.pnas.org/lookup/doi/10.1073/pnas.1421675112>.
- , and Coauthors, 2015c: 21st century United States emissions mitigation could increase water stress more than the climate change it is mitigating. *Proc. Natl. Acad. Sci.*, **112**, 10635–10640, doi:10.1073/pnas.1421675112.
- , and Coauthors, 2015d: 21st century United States emissions mitigation could increase water stress more than the climate change it is mitigating. *Proc. Natl. Acad. Sci.*, **112**, 10635–10640, doi:10.1073/pnas.1421675112. <http://www.pnas.org/lookup/doi/10.1073/pnas.1421675112> (Accessed November 11, 2018).
- Helliwell, D. R., 1969: Valuation of wildlife resources. *Reg. Stud.*, **3**, 41–47, doi:10.1080/09595236900185051. <http://www.tandfonline.com/doi/abs/10.1080/09595236900185051> (Accessed April 5, 2019).
- Hellmann, J. J., J. E. Byers, B. G. Bierwagen, and J. S. Dukes, 2008: Five Potential Consequences of Climate Change for Invasive Species. *Conserv. Biol.*, **22**, 534–543, doi:10.1111/j.1523-1739.2008.00951.x. <http://doi.wiley.com/10.1111/j.1523-1739.2008.00951.x> (Accessed April 4, 2019).
- Henderson, B. B. *et al.* Greenhouse gas mitigation potential of the world's grazing lands: Modeling soil carbon and nitrogen fluxes of mitigation practices. *Agric. Ecosyst. Environ.* **207**, 91–100 (2015).
- Hendrickson, D., C. Smith, and N. Eikenberry, 2006: Fruit and vegetable access in four low-income food deserts communities in Minnesota. *Agric. Human Values*, **23**, 371–383, doi:10.1007/s10460-006-9002-8. <http://link.springer.com/10.1007/s10460-006-9002-8> (Accessed April 15, 2019).
- Hengsdijk, H., and W. J. de Boer, 2017: Post-harvest management and post-harvest losses of cereals in Ethiopia. *Food Secur.*, **9**, 945–958, doi:10.1007/s12571-017-0714-y.

<http://link.springer.com/10.1007/s12571-017-0714-y> (Accessed May 10, 2018).

- Henson, S., and R. Loader, 2001: Barriers to Agricultural Exports from Developing Countries: The Role of Sanitary and Phytosanitary Requirements. *World Dev.*, **29**, 85–102, doi:10.1016/S0305-750X(00)00085-1. <https://www.sciencedirect.com/science/article/abs/pii/S0305750X00000851> (Accessed April 15, 2019).
- Herder, M.; Moreno, G.; Mosquera-Losada, R.M.; Palma, J.H.N.; Sidiropoulou, A.; Freijanes, J.J.S.; CrousDuran, J.; Paulo, J.A.; Tomé, M.; Pantera, A.; Papanastasis, V.P.; Mantzanas, K.; Pachana, P.; Papadopoulos, A.; Plieninger, T. & Burgess, P.J. 2017. Current extent and stratification of agroforestry in the European Union. [Agriculture, Ecosystems & Environment](#) **241**, 121-132.
- Hernandez-Morcillo, M., P. Burgess, J. Mirck, A. Pantera, and T. Plieninger, 2018: Scanning agroforestry-based solutions for climate change mitigation and adaptation in Europe. *Environ. Sci. Policy*, **80**, 44–52. <https://www.sciencedirect.com/science/article/pii/S1462901117310250> (Accessed November 11, 2018).
- Herrero, M., B. Henderson, P. Havlík, ... P. T.-N. C., and undefined 2016, Greenhouse gas mitigation potentials in the livestock sector. *nature.com*,. <https://www.nature.com/articles/nclimate2925> (Accessed November 8, 2018).
- Herrero, M., and Coauthors, 2016: Greenhouse gas mitigation potentials in the livestock sector. *Nat. Clim. Chang.*, **6**, 452–461, doi:10.1038/nclimate2925. <https://www.nature.com/nclimate/journal/v6/n5/abs/nclimate2925.html> (Accessed December 20, 2017).
- Herrmann, S., and C. Hutchinson, 2005: The changing contexts of the desertification debate. *J. Arid Environ.*, **63**, 538–555. <https://www.sciencedirect.com/science/article/pii/S0140196305000492> (Accessed November 11, 2018).
- Hertel, T. W., 2015: The challenges of sustainably feeding a growing planet. *Food Secur.*, **7**, 185–198, doi:10.1007/s12571-015-0440-2.
- , and U. L. C. Baldos, 2016: Attaining food and environmental security in an era of globalization. *Glob. Environ. Chang.*, **41**, 195–205, doi:10.1016/J.GLOENVCHA.2016.10.006. <https://www.sciencedirect.com/science/article/abs/pii/S0959378016304046> (Accessed April 15, 2019).
- Hill, R., J. Davies, I. C. Bohnet, C. J. Robinson, K. Maclean, and P. L. Pert, 2015: Collaboration mobilises institutions with scale-dependent comparative advantage in landscape-scale biodiversity conservation. *Environ. Sci. Policy*, **51**, 267–277, doi:10.1016/J.ENVSCL.2015.04.014. <https://www.sciencedirect.com/science/article/pii/S1462901115000878> (Accessed April 4, 2019).
- Hillbruner, C., and G. Moloney, 2012: When early warning is not enough—Lessons learned from the 2011 Somalia Famine. *Glob. Food Sec.*, **1**, 20–28, doi:10.1016/J.GFS.2012.08.001. <https://www.sciencedirect.com/science/article/pii/S2211912412000107> (Accessed June 1, 2018).
- Hine, J. L., 1993: *Transport and marketing priorities to improve food security in Ghana and the rest of Africa*.
- Hinkel, J., and Coauthors, 2014: Coastal flood damage and adaptation costs under 21st century sea-level rise. *Proc. Natl. Acad. Sci.*, **111**, 3292–3297. <http://www.pnas.org/content/early/2014/01/29/1222469111.short> (Accessed November 11, 2018).
- HLPE, 2017: *Nutrition and food systems. A report by the High Level Panel of Experts on Food Security and Nutrition of the Committee on World Food Security*. Rome, <http://www.fao.org/cfs/cfs->

hlpe/reports/report-12-elaboration-process/en/.

- Hodges, R. J., J. C. Buzby, and B. Bennett, 2011: Postharvest losses and waste in developed and less developed countries: Opportunities to improve resource use. *J. Agric. Sci.*, doi:10.1017/S0021859610000936.
- Hof, C., A. Voskamp, M. F. Biber, K. Böhning-Gaese, E. K. Engelhardt, A. Niamir, S. G. Willis, and T. Hickler, 2018: Bioenergy cropland expansion may offset positive effects of climate change mitigation for global vertebrate diversity. *Proc. Natl. Acad. Sci.*, **115**, 13294 LP-13299, doi:10.1073/pnas.1807745115. <http://www.pnas.org/content/115/52/13294.abstract>.
- Hoff, H., and L. M. Bouwer, 2003: *Risk Management in Water and Climate-the Role of Insurance and Other Financial Services Adaptation to climate change View project Cities and Flood Management View project*. <https://www.researchgate.net/publication/241237030> (Accessed April 16, 2019).
- Hoffmann, T., and Coauthors, 2013: Humans and the missing C-sink: erosion and burial of soil carbon through time. *Earth Surf. Dyn.*, **1**, 45–52. <https://www.earth-surf-dynam.net/1/45/2013/esurf-1-45-2013.html> (Accessed November 10, 2018).
- Hooijer, A., S. Page, J. G. Canadell, M. Silvius, J. Kwadijk, H. Wosten, and J. Jauhiainen, 2010: Current and future CO₂ emissions from drained peatlands in Southeast Asia. *Biogeosciences*, **7**, 1505–1514, doi:10.5194/bg-7-1505-2010.
- Hosonuma, N., M. Herold, V. De Sy, R. S. De Fries, M. Brockhaus, L. Verchot, A. Angelsen, and E. Romijn, 2012: An assessment of deforestation and forest degradation drivers in developing countries. *Environ. Res. Lett.*, **7**, 44009, doi:10.1088/1748-9326/7/4/044009. <http://stacks.iop.org/1748-9326/7/i=4/a=044009?key=crossref.1a00aa77eac35c904bf7e007011d4763> (Accessed November 11, 2018).
- Hou, S., A. Li, B. Han, and P. Zhou, 2013: An Early Warning System for Regional Rain-Induced Landslide Hazard. *Int. J. Geosci.*, **4**, 584–587, doi:10.4236/ijg.2013.43053. <http://dx>. (Accessed April 16, 2019).
- Houghton, R. A. & Nassikas, A. A. Negative emissions from stopping deforestation and forest degradation, globally. *Glob. Chang. Biol.* **24**, 350–359 (2018).
- Houghton, R. A., B. Byers, and A. A. Nassikas, 2015: A role for tropical forests in stabilizing atmospheric CO₂. *Nat. Clim. Chang.*, **5**, 1022–1023, doi:10.1038/nclimate2869.
- Howard, J. *et al.* Clarifying the role of coastal and marine systems in climate mitigation. *Front. Ecol. Environ.* (2017).
- Howard, P. H., 2015: Intellectual Property and Consolidation in the Seed Industry. *Crop Sci.*, **55**, 2489, doi:10.2135/cropsci2014.09.0669. <https://dl.sciencesocieties.org/publications/cs/abstracts/55/6/2489> (Accessed June 1, 2018).
- Howarth, C., and I. Monasterolo, 2017: Opportunities for knowledge co-production across the energy-food-water nexus: Making interdisciplinary approaches work for better climate decision making. *Environ. Sci. Policy*, **75**, 103–110, doi:10.1016/j.envsci.2017.05.019.
- Howell, T. A., S. R. Evett, J. A. Tolk, K. S. Copeland, and T. H. Marek, 2015: Evapotranspiration, water productivity and crop coefficients for irrigated sunflower in the U.S. Southern High Plains. *Agric. Water Manag.*, **162**, 33–46, doi:10.1016/J.AGWAT.2015.08.008. <https://www.sciencedirect.com/science/article/pii/S0378377415300743> (Accessed April 12, 2019).
- Hawken, P. *Project Drawdown: The most comprehensive plan ever proposed to reverse global warming*. (Penguin Books, 2017).

- Huang, J., H. Yu, X. Guan, G. Wang, and R. Guo, 2016: Accelerated dryland expansion under climate change. *Nat. Clim. Chang.*, **6**, 166–171, doi:10.1038/nclimate2837.
- Hudiburg, T. W., S. C. Davis, W. Parton, and E. H. Delucia, 2015: Bioenergy crop greenhouse gas mitigation potential under a range of management practices. *GCB Bioenergy*, doi:10.1111/gcbb.12152.
- Humpenöder, F., and Coauthors, 2014a: Investigating afforestation and bioenergy CCS as climate change mitigation strategies. *Environ. Res. Lett.*, **9**, 64029, doi:10.1088/1748-9326/9/6/064029. <http://stacks.iop.org/1748-9326/9/i=6/a=064029?key=crossref.5fa44a1462d2acebebaa002315d8e4a6> (Accessed May 30, 2018).
- , and Coauthors, 2014b: Investigating afforestation and bioenergy CCS as climate change mitigation strategies. *Environ. Res. Lett.*, **9**, 64029, doi:10.1088/1748-9326/9/6/064029. <http://stacks.iop.org/1748-9326/9/i=6/a=064029?key=crossref.5fa44a1462d2acebebaa002315d8e4a6> (Accessed May 30, 2018).
- , and Coauthors, 2018a: Large-scale bioenergy production: how to resolve sustainability trade-offs? *Environ. Res. Lett.*, **13**, 24011, doi:10.1088/1748-9326/aa9e3b. <http://stacks.iop.org/1748-9326/13/i=2/a=024011?key=crossref.ecb4dde273241bb18b2b44b439e5f874>.
- , and Coauthors, 2018b: Large-scale bioenergy production: how to resolve sustainability trade-offs? *Environ. Res. Lett.*, **13**, 24011, doi:10.1088/1748-9326/aa9e3b. <http://stacks.iop.org/1748-9326/13/i=2/a=024011?key=crossref.ecb4dde273241bb18b2b44b439e5f874> (Accessed November 11, 2018).
- Huntjens, P., L. Lebel, C. Pahl-Wostl, J. Camkin, R. Schulze, and N. Kranz, 2012: Institutional design propositions for the governance of adaptation to climate change in the water sector. *Glob. Environ. Chang.*, **22**, 67–81, doi:10.1016/j.gloenvcha.2011.09.015.
- Huppmann, D., and Coauthors, 2018: IAMC 1.5 C Scenario Explorer and Data hosted by IIASA. *Integrated Assessment Modeling Consortium & International Institute for Applied Systems Analysis data.ene.iiasa.ac.at/iamc-1.5c-explorer*.
- Hurteau, M. D., J. B. Bradford, P. Z. Fulé, A. H. Taylor, and K. L. Martin, 2014: Climate change, fire management, and ecological services in the southwestern US. *For. Ecol. Manage.*, **327**, 280–289, doi:10.1016/J.FORECO.2013.08.007. <https://www.sciencedirect.com/science/article/pii/S0378112713005343> (Accessed May 31, 2018).
- Idris Medugu, N., M. Rafee Majid, F. Johar, and I. D. Choji, 2010a: The role of afforestation programme in combating desertification in Nigeria. *Int. J. Clim. Chang. Strateg. Manag.*, **2**, 35–47, doi:10.1108/17568691011020247. <http://www.emeraldinsight.com/doi/10.1108/17568691011020247> (Accessed November 11, 2018).
- , ———, ———, and ———, 2010b: The role of afforestation programme in combating desertification in Nigeria. *Int. J. Clim. Chang. Strateg. Manag.*, **2**, 35–47, doi:10.1108/17568691011020247. <http://www.emeraldinsight.com/doi/10.1108/17568691011020247> (Accessed April 16, 2019).
- , ———, ———, and ———, 2010c: The role of afforestation programme in combating desertification in Nigeria. *Int. J. Clim. Chang. Strateg. Manag.*, **2**, 35–47, doi:10.1108/17568691011020247. <http://www.emeraldinsight.com/doi/10.1108/17568691011020247> (Accessed May 30, 2018).
- Iglesias, A., L. Garrote, F. Flores, and M. Moneo, 2007: Challenges to Manage the Risk of Water Scarcity and Climate Change in the Mediterranean. *Water Resour. Manag.*, **21**, 775–788, doi:10.1007/s11269-006-9111-6. <http://link.springer.com/10.1007/s11269-006-9111-6> (Accessed

April 16, 2019).

Iho, A., M. Ribaudó, and K. Hyytiäinen, 2015: Water protection in the Baltic Sea and the Chesapeake Bay: Institutions, policies and efficiency. *Mar. Pollut. Bull.*, **93**, 81–93, doi:10.1016/J.MARPOLBUL.2015.02.011.

<https://www.sciencedirect.com/science/article/pii/S0025326X15000909> (Accessed April 16, 2019).

Immerzeel, D. J., P. A. Verweij, F. van der Hilst, and A. P. C. Faaij, 2014: Biodiversity impacts of bioenergy crop production: a state-of-the-art review. *GCB Bioenergy*, **6**, 183–209, doi:10.1111/gcbb.12067. <https://doi.org/10.1111/gcbb.12067>.

Ingram, J., 2011: A food systems approach to researching food security and its interactions with global environmental change. *Food Secur.*, doi:DOI 10.1007/s12571-011-0149-9.

Ingram, J., R. Dyball, M. Howden, S. Vermeulen, T. Ganett, B. Redlingshöfer, S. Guilbert, and J. Porter, 2016a: Food security, food systems, and environmental change. *Solut. J.*, **7**, 2154–0926.

———, R. Dyball, M. Howden, S. Vermeulen, T. Ganett, B. Redlingshöfer, S. Guilbert, and J. Porter, 2016b: *Food security, food systems, and environmental change*.

Jordan, C.-M., X. Hu, A. Arvesen, P. Kauppi, and F. Cherubini, 2018: Contribution of forest wood products to negative emissions: historical comparative analysis from 1960 to 2015 in Norway, Sweden and Finland. *Carbon Balance Manag.*, **13**, 12, doi:10.1186/s13021-018-0101-9. <https://cbmjournal.biomedcentral.com/articles/10.1186/s13021-018-0101-9> (Accessed April 19, 2019).

IPBES, 2018: *Summary for Policymakers of the Thematic assessment Report on Land Degradation and Restoration of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services*. R. Scholes et al., Eds. IPBES Secretariat, Bonn, Germany, 1-31 pp.

IPBES (2018): The IPBES assessment report on land degradation and restoration. Montanarella, L., Scholes, R., and Brainich, A. (eds.). Secretariat of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services, Bonn, Germany. 744 pages.

IPCC, 2000: *Land Use, Land-Use Change and Forestry*. R.T. Watson, I.R. Noble, B. Bolin, N.H. Ravindranath, D.J. Verardo, and D.J. Dokken, Eds. Cambridge University Press.

———, 2014: *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. O. Edenhofer et al., Eds. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.

———, 2018: *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of 1.5 °C above pre-industrial levels and related global greenhouse gas emission pathways, in the context of strengthening the global response to the threat of climate change*. <http://www.ipcc.ch/report/sr15/>.

IPCC SR15, *Global Warming of 1.5°C: A Special Report of Working Groups I, II and III of the Intergovernmental Panel on Climate Change*.

IPCC (2011) *Bioenergy IN IPCC Special Report on Renewable Energy Sources and Climate Change Mitigation*. Cambridge, United Kingdom and New York, NY, USA, Cambridge University Press

IPCC (2006) **2006 IPCC Guidelines for National Greenhouse Gas Inventories**, Prepared by the National Greenhouse Gas Inventories Programme,

Eggleston H.S., Buendia L., Miwa K., Ngara T. and Tanabe K. (eds). Published: IGES, Japan

ITPS-FAO, 2015: *Status of the World's Soil Resources (SWSR)*. Rome, <http://www.fao.org/3/a-i5199e.pdf> (Accessed June 5, 2018).

Ivanic, M., and W. Martin, 2008: Implications of higher global food prices for poverty in low-income countries ¹. *Agric. Econ.*, **39**, 405–416, doi:10.1111/j.1574-0862.2008.00347.x. <http://doi.wiley.com/10.1111/j.1574-0862.2008.00347.x> (Accessed November 11, 2018).

Iyer, G., and Coauthors, 2018: Implications of sustainable development considerations for comparability across nationally determined contributions. *Nat. Clim. Chang.*, **8**, 124–129, doi:10.1038/s41558-017-0039-z.

Jabareen, Y. R., 2006: Sustainable Urban Forms. *J. Plan. Educ. Res.*, **26**, 38–52, doi:10.1177/0739456X05285119. <http://journals.sagepub.com/doi/10.1177/0739456X05285119> (Accessed April 16, 2019).

Jacinthe, P. A., and R. Lal, 2001: A mass balance approach to assess carbon dioxide evolution during erosional events. *L. Degrad. Dev.*, **12**, 329–339, doi:10.1002/ldr.454. <http://doi.wiley.com/10.1002/ldr.454> (Accessed November 11, 2018).

Jacoby, H. G., and E. Skoufias, 1997: Risk, Financial Markets, and Human Capital in a Developing Country. *Rev. Econ. Stud.*, **64**, 311, doi:10.2307/2971716. <https://academic.oup.com/restud/article-lookup/doi/10.2307/2971716> (Accessed April 16, 2019).

Jactel, H., and Coauthors, 2017: Tree Diversity Drives Forest Stand Resistance to Natural Disturbances. *Curr. For. Reports*, **3**, 223–243, doi:10.1007/s40725-017-0064-1.

James, S. J., and C. James, 2010: The food cold-chain and climate change. *Food Res. Int.*, **43**, 1944–1956, doi:10.1016/j.foodres.2010.02.001.

Jans, Y., G. Berndes, J. Heinke, W. Lucht, and D. Gerten, 2018: Biomass production in plantations: Land constraints increase dependency on irrigation water. *GCB Bioenergy*, **10**, 628–644, doi:10.1111/gcbb.12530.

Jantz, P., S. Goetz, and N. Laporte, 2014: Carbon stock corridors to mitigate climate change and promote biodiversity in the tropics. *Nat. Clim. Chang.*, **4**, 138–142, doi:10.1038/nclimate2105. <http://www.nature.com/articles/nclimate2105> (Accessed June 6, 2018).

Jargowsky, P., 2002: Sprawl, concentration of poverty, and urban inequality. *Urban Sprawl Causes, Consequences and Policy Responses*, Washington D.C., 39–72.

Jat, H. S., G. Singh, R. Singh, M. Choudhary, M. L. Jat, M. K. Gathala, and D. K. Sharma, 2015: Management influence on maize-wheat system performance, water productivity and soil biology. *Soil Use Manag.*, **31**, 534–543, doi:10.1111/sum.12208. <http://doi.wiley.com/10.1111/sum.12208> (Accessed April 12, 2019).

Jat, M., J. Dagar, T. Sapkota, ... B. G.-A. in, and undefined 2016, Climate change and agriculture: adaptation strategies and mitigation opportunities for food security in South Asia and Latin America. *Elsevier*.

—, and Coauthors, 2016: Climate change and agriculture: adaptation strategies and mitigation opportunities for food security in South Asia and Latin America. *Adv. Agron.*, **137**, 127–235. <https://www.sciencedirect.com/science/article/pii/S0065211315300055> (Accessed November 10, 2018).

- Jauhainen, J., S. Limin, H. Silvennoinen, and H. Vasander, 2008: Carbon dioxide and methane fluxes in drained tropical peat before and after hydrological restoration. *Ecology*, **89**, 3503–3514, doi:10.1890/07-2038.1. <http://doi.wiley.com/10.1890/07-2038.1> (Accessed June 1, 2018).
- Jaworski, A., 2016: Encouraging Climate Adaptation through Reform of Federal Crop Insurance Subsidies. *New York Univ. Law Rev.*, **91**, 1684. <https://heinonline.org/HOL/Page?handle=hein.journals/nylr91&id=1721&div=&collection=> (Accessed June 1, 2018).
- Jebli, M., and S. Youssef, 2017: The role of renewable energy and agriculture in reducing CO2 emissions: Evidence for North Africa countries. *Ecol. Indic.*, **74**, 295–301. <https://www.sciencedirect.com/science/article/pii/S1470160X16306690> (Accessed November 11, 2018).
- Jeffery, S., D. Abalos, M. Prodana, A. C. Bastos, J. W. van Groenigen, B. A. Hungate, and F. Verheijen, 2017: Biochar boosts tropical but not temperate crop yields. *Environ. Res. Lett.*, **12**, 53001, doi:10.1088/1748-9326/aa67bd. <http://stacks.iop.org/1748-9326/12/i=5/a=053001?key=crossref.bf9e05e5fb214cb193e8f1c2d66a7ac9> (Accessed June 6, 2018).
- Jenks, M. (Michael), and R. Burgess, 2000: *Compact cities : sustainable urban forms for developing countries*. E. & F.N. Spon, 356 pp. https://books.google.co.nz/books?hl=en&lr=&id=glqRAGAAQBAJ&oi=fnd&pg=PP1&dq=Compact+cities:+sustainable+urban+forms+for+developing+countries,+&ots=qRM7Jr3gi7&sig=CEMvH0bRjpYWeAB0m_x2v5xhn0U#v=onepage&q=Compact+cities%3A+sustainable+urban+forms+for+developing+countries%2C&f=false (Accessed April 16, 2019).
- Jiang, Y., 2009: China's water scarcity. *J. Environ. Manage.*, **90**, 3185–3196, doi:10.1016/J.JENVMAN.2009.04.016. <https://www.sciencedirect.com/science/article/pii/S0301479709001339> (Accessed April 16, 2019).
- , 2015: China's water security: Current status, emerging challenges and future prospects. *Environ. Sci. Policy*, **54**, 106–125, doi:10.1016/j.envsci.2015.06.006. <http://dx.doi.org/10.1016/j.envsci.2015.06.006>.
- Johnson, F. X., and S. Silveira, 2014: Pioneer countries in the transition to alternative transport fuels: Comparison of ethanol programmes and policies in Brazil, Malawi and Sweden. *Environ. Innov. Soc. Transitions*, **11**, 1–24, doi:10.1016/j.eist.2013.08.001.
- Johnston, F. H., and Coauthors, 2012: Estimated global mortality attributable to smoke from landscape fires. *Environ. Health Perspect.*, doi:10.1289/ehp.1104422.
- Jones, A. D., and Coauthors, 2013: Greenhouse gas policy influences climate via direct effects of land-use change. *J. Clim.*, **26**, 3657–3670, doi:10.1175/JCLI-D-12-00377.1.
- , K. V. Calvin, W. D. Collins, and J. Edmonds, 2015: Accounting for radiative forcing from albedo change in future global land-use scenarios. *Clim. Change*, **131**, 691–703, doi:10.1007/s10584-015-1411-5.
- Jones, H. P., D. G. Hole, and E. S. Zavaleta, 2012: Harnessing nature to help people adapt to climate change. *Nat. Clim. Chang.*, **2**, 504–509, doi:10.1038/nclimate1463. <http://www.nature.com/doifinder/10.1038/nclimate1463> (Accessed June 1, 2018).
- Jongwanich, J., 2009: The impact of food safety standards on processed food exports from developing countries. *Food Policy*, **34**, 447–457, doi:10.1016/J.FOODPOL.2009.05.004. <https://www.sciencedirect.com/science/article/abs/pii/S0306919209000529> (Accessed April 15, 2019).

- Joosten, H., and J. Couwenberg, 2008: Peatlands and carbon. *Assess. peatlands, Biodivers. Clim. Chang. Glob. Environ. Centre, Kuala Lumpur Wetl. Int. Wageningen*, 99–117.
- Kadiyala, S., J. Harris, D. Headey, S. Yosef, and S. Gillespie, 2014: Agriculture and nutrition in India: mapping evidence to pathways. *Ann. N. Y. Acad. Sci.*, **1331**, 43–56, doi:10.1111/nyas.12477. <http://www.rchiips.org/> (Accessed April 15, 2019).
- Kaminski, J., and L. Christiaensen, 2014: *Post-Harvest Loss in Sub-Saharan Africa — What Do Farmers Say?* The World Bank, <http://elibrary.worldbank.org/doi/book/10.1596/1813-9450-6831> (Accessed April 15, 2019).
- Kantor, L., 2001: Community food security programs improve food access. *Food Rev.*, **24**, 20–26.
- Kapos, V., and Coauthors, 2008: *Carbon and biodiversity: a demonstration atlas*. Cambridge, UK, 163–97 pp.
- Karim, M. R., and A. Thiel, 2017: Role of community based local institution for climate change adaptation in the Teesta riverine area of Bangladesh. *Clim. Risk Manag.*, **17**, 92–103, doi:10.1016/j.crm.2017.06.002. <https://linkinghub.elsevier.com/retrieve/pii/S2212096317300177> (Accessed November 7, 2018).
- Karjalainen, E., T. Sarjala, and H. Raitio, 2009: Promoting human health through forests: overview and major challenges. *Environ. Health Prev. Med.*, **15**, 1, doi:10.1007/s12199-008-0069-2.
- Katz, B. G., 1989: Influence of Mineral Weathering Reactions on the Chemical-Composition of Soil-Water, Springs, and Ground-Water, Catocin Mountains, Maryland. *Hydrol. Process.*, **3**, 185–202.
- P. Kauppi, R. Sedjo, M. Apps, C. Cerri, T. Fujimoro, H. Janzen, O. Krankina, W. Makundi, G. Marland, O. Masera, G.-J. Nabuurs, W. Razali, N.H. Ravindranath **Technological and economic potential of options to enhance, maintain and manage biological carbon reservoirs and geo-engineering**
- B. Metz, O. Davidson, R. Swart, J. Pan (Eds.), *Climate Change 2001-Mitigation*, Cambridge Univ. Press, UK (2001), pp. 301-343
- Kauppi, P. E., V. Sandström, and A. Lipponen, 2018: Forest resources of nations in relation to human well-being. *PLoS One*, **13**, e0196248, doi:10.1371/journal.pone.0196248. <http://dx.plos.org/10.1371/journal.pone.0196248> (Accessed June 2, 2018).
- Keddy, P., L. Fraser, A. Solomeshch, and W. Junk, 2009: Wet and wonderful: the world's largest wetlands are conservation priorities. *Bioscience*, **59**, 39–51. <https://academic.oup.com/bioscience/article-abstract/59/1/39/306994> (Accessed November 11, 2018).
- Keding, G. B., K. Schneider, and I. Jordan, 2013: Production and processing of foods as core aspects of nutrition-sensitive agriculture and sustainable diets. *Food Secur.*, **5**, 825–846, doi:10.1007/s12571-013-0312-6. <http://link.springer.com/10.1007/s12571-013-0312-6> (Accessed April 16, 2019).
- Keenan, R. J., G. A. Reams, F. Achard, J. V. de Freitas, A. Grainger, and E. Lindquist, 2015: Dynamics of global forest area: Results from the FAO Global Forest Resources Assessment 2015. *For. Ecol. Manage.*, **352**, 9–20, doi:10.1016/J.FORECO.2015.06.014. <https://www.sciencedirect.com/science/article/pii/S0378112715003400> (Accessed April 5, 2019).
- Keesstra, S., J. Bouma, J. Wallinga, P. Tittonell, P. S.- Soil, and undefined 2016, The significance of soils and soil science towards realization of the United Nations Sustainable Development Goals. *soil-journal.net*, <https://www.soil-journal.net/2/111/2016/> (Accessed November 10, 2018).
- Keesstra, S., J. Nunes, A. Novara, D. Finger, D. Avelar, Z. Kalantari, and A. Cerdà, 2018: The superior

- effect of nature based solutions in land management for enhancing ecosystem services. *Sci. Total Environ.*, **610–611**, 997–1009. <https://www.sciencedirect.com/science/article/pii/S0048969717320752> (Accessed June 2, 2018).
- Kemper, J., 2015a: Biomass and carbon dioxide capture and storage: A review. *Int. J. Greenh. Gas Control*, **40**, 401–430, doi:10.1016/j.ijggc.2015.06.012. <http://linkinghub.elsevier.com/retrieve/pii/S1750583615002650>.
- , 2015b: Biomass and carbon dioxide capture and storage: A review. *Int. J. Greenh. Gas Control*, **40**, 401–430, doi:10.1016/j.ijggc.2015.06.012. <https://linkinghub.elsevier.com/retrieve/pii/S1750583615002650> (Accessed April 18, 2019).
- Kerr, R. B., 2005: Food Security in Northern Malawi: Gender, Kinship Relations and Entitlements in Historical Context. *J. South. Afr. Stud.*, **31**, 53–74, doi:10.1080/03057070500035679. <http://www.tandfonline.com/doi/abs/10.1080/03057070500035679> (Accessed April 16, 2019).
- Khan, M. M., N. B. Mock, and W. B. Bertrand, 1992: Composite Indicators for Famine Early Warning Systems. *Disasters*, **16**, 195–206, doi:10.1111/j.1467-7717.1992.tb00398.x. <http://doi.wiley.com/10.1111/j.1467-7717.1992.tb00398.x> (Accessed April 16, 2019).
- Kim, S. H., and Coauthors, 2016a: Balancing global water availability and use at basin scale in an integrated assessment model. *Clim. Change*, doi:10.1007/s10584-016-1604-6.
- , and Coauthors, 2016b: Balancing global water availability and use at basin scale in an integrated assessment model. *Clim. Change*, **136**, 217–231, doi:10.1007/s10584-016-1604-6. <http://link.springer.com/10.1007/s10584-016-1604-6> (Accessed November 11, 2018).
- Kim Pingoud, Tommi Ekholm, Risto Sievänen, Saija Huuskonen, Jari Hynynen, Trade-offs between forest carbon stocks and harvests in a steady state – A multi-criteria analysis, *Journal of Environmental Management*, Volume 210, 2018, Pages 96-103
- Kindermann, G., and Coauthors, 2008: Global cost estimates of reducing carbon emissions through avoided deforestation. *Proc. Natl. Acad. Sci.*, **105**, 10302–10307, doi:10.1073/pnas.0710616105. <http://www.pnas.org/cgi/doi/10.1073/pnas.0710616105> (Accessed November 11, 2018).
- King, R. T., 1966: Wildlife and man. *New York Conserv.*, **20**, 8–11.
- Kissinger, M., C. Sussmann, and C. Dorward, 2018: Local or global: A biophysical analysis of a regional food system. *Food Syst.*, <https://www.cambridge.org/core/journals/renewable-agriculture-and-food-systems/article/local-or-global-a-biophysical-analysis-of-a-regional-food-system/BEB1826C608FC06CE6CD221D25C1EA6D> (Accessed May 10, 2018).
- Kline, K. L., and Coauthors, 2017: Reconciling food security and bioenergy: priorities for action. *GCB Bioenergy*, doi:10.1111/gcbb.12366.
- Kloppenborg, J., 2010: Impeding Dispossession, Enabling Repossession: Biological Open Source and the Recovery of Seed Sovereignty. *J. Agrar. Chang.*, **10**, 367–388, doi:10.1111/j.1471-0366.2010.00275.x. <http://doi.wiley.com/10.1111/j.1471-0366.2010.00275.x> (Accessed June 1, 2018).
- Kloppenborg, J., 2014: Re-purposing the master’s tools: the open source seed initiative and the struggle for seed sovereignty. *J. Peasant Stud.*, **41**, 1225–1246, doi:10.1080/03066150.2013.875897. <http://www.tandfonline.com/doi/abs/10.1080/03066150.2013.875897> (Accessed April 16, 2019).
- Knoke, T., and Coauthors, 2014: Afforestation or intense pasturing improve the ecological and economic value of abandoned tropical farmlands. *Nat. Commun.*, doi:10.1038/ncomms6612.
- Knorr, W., A. Arneth, and L. Jiang, 2016: Demographic controls of future global fire risk. *Nat. Clim. Chang.*, **6**, 781–785, doi:10.1038/nclimate2999.

- Kobayashi, Y., and A. S. Mori, 2017: The Potential Role of Tree Diversity in Reducing Shallow Landslide Risk. *Environ. Manage.*, **59**, 807–815, doi:10.1007/s00267-017-0820-9.
- Kombe, W. J., 2005: Land use dynamics in peri-urban areas and their implications on the urban growth and form: the case of Dar es Salaam, Tanzania. *Habitat Int.*, **29**, 113–135, doi:10.1016/S0197-3975(03)00076-6. <https://www.sciencedirect.com/science/article/pii/S0197397503000766> (Accessed April 16, 2019).
- Kondylis, F., V. Mueller, G. Sheriff, and S. Zhu, 2016: Do female instructors reduce gender bias in diffusion of sustainable land management techniques? Experimental evidence from Mozambique. *World Dev.*, **78**, 436–449. <https://www.sciencedirect.com/science/article/pii/S0305750X15002582> (Accessed November 10, 2018).
- Kongsager, R., B. Locatelli, and F. Chazarin, 2016: Addressing Climate Change Mitigation and Adaptation Together: A Global Assessment of Agriculture and Forestry Projects. *Environ. Manage.*, **57**, 271–282, doi:10.1007/s00267-015-0605-y. <http://link.springer.com/10.1007/s00267-015-0605-y> (Accessed June 3, 2018).
- Kowalski, J., and T. Conway, 2018: Branching out: The inclusion of urban food trees in Canadian urban forest management plans. *Urban For. Urban Green.*, <https://www.sciencedirect.com/science/article/pii/S1618866718300736> (Accessed November 11, 2018).
- Kramer, R. A., W. T. McSweeney, and R. W. Stavros, 1983: Soil Conservation with Uncertain Revenues and Input Supplies. *Am. J. Agric. Econ.*, **65**, 694, doi:10.2307/1240457. <https://academic.oup.com/ajae/article-lookup/doi/10.2307/1240457> (Accessed April 16, 2019).
- Krause, A., and Coauthors, 2017: Global consequences of afforestation and bioenergy cultivation on ecosystem service indicators. *Biogeosciences*, doi:10.5194/bg-14-4829-2017.
- Kraxner, F., and Coauthors, 2013: Global bioenergy scenarios - Future forest development, land-use implications, and trade-offs. *Biomass and Bioenergy*, doi:10.1016/j.biombioe.2013.02.003.
- Kreidenweis, U., F. Humpenöder, M. Stevanović, B. L. Bodirsky, E. Kriegler, H. Lotze-Campen, and A. Popp, 2016a: Afforestation to mitigate climate change: impacts on food prices under consideration of albedo effects. *Environ. Res. Lett.*, **11**, 85001, doi:10.1088/1748-9326/11/8/085001. <http://stacks.iop.org/1748-9326/11/i=8/a=085001?key=crossref.498dab12c59b27f71805e8cdbafc36f1> (Accessed November 9, 2018).
- Kriegler, E., and Coauthors, 2014: The role of technology for achieving climate policy objectives: overview of the EMF 27 study on global technology and climate policy strategies. *Clim. Change*, **123**, 353–367, doi:10.1007/s10584-013-0953-7. <https://doi.org/10.1007/s10584-013-0953-7>.
- Kriegler, E., and Coauthors, 2017: Fossil-fueled development (SSP5): An energy and resource intensive scenario for the 21st century. *Glob. Environ. Chang.*, **42**, doi:10.1016/j.gloenvcha.2016.05.015.
- Kriegler, E., and Coauthors, 2018a: Short term policies to keep the door open for Paris climate goals. *Environ. Res. Lett.*, **13**, 74022, doi:10.1088/1748-9326/aac4f1.
- , and Coauthors, 2018b: Pathways limiting warming to 1.5°C: a tale of turning around in no time? *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, **376**, 20160457, doi:10.1098/rsta.2016.0457.
- Kumar, D., and P. Kalita, 2017: Reducing Postharvest Losses during Storage of Grain Crops to Strengthen Food Security in Developing Countries. *Foods*, doi:10.3390/foods6010008.
- Kummu, M., H. De Moel, M. Porkka, ... S. S.-S. of the total, and undefined 2012, Lost food, wasted resources: Global food supply chain losses and their impacts on freshwater, cropland, and fertiliser

- use. *Elsevier*,. <https://www.sciencedirect.com/science/article/pii/S0048969712011862> (Accessed April 18, 2019).
- Kummu, M., P. J. Ward, H. de Moel, and O. Varis, 2010: Is physical water scarcity a new phenomenon? Global assessment of water shortage over the last two millennia. *Environ. Res. Lett.*, **5**, 34006, doi:10.1088/1748-9326/5/3/034006. <http://stacks.iop.org/1748-9326/5/i=3/a=034006?key=crossref.273693a90fc59f2efed1721b344c6ce6> (Accessed November 11, 2018).
- Kummu, M., H. de Moel, M. Porkka, S. Siebert, O. Varis, and P. J. Ward, 2012: Lost food, wasted resources: Global food supply chain losses and their impacts on freshwater, cropland, and fertiliser use. *Sci. Total Environ.*, **438**, 477–489, doi:10.1016/J.SCITOTENV.2012.08.092. <https://www.sciencedirect.com/science/article/pii/S0048969712011862> (Accessed April 5, 2019).
- Kurian, M., 2017: The water-energy-food nexus: trade-offs, thresholds and transdisciplinary approaches to sustainable development. *Environ. Sci. Policy*,. <https://www.sciencedirect.com/science/article/pii/S1462901116305184> (Accessed May 6, 2018).
- Kurz, W., C. Smyth, and T. Lemprière, 2016: Climate change mitigation through forest sector activities: Principles, potential and priorities 1. *Unasylva*, **67**, 61–67.
- Kyle, P., C. Müller, K. Calvin, and A. Thomson, 2014: Meeting the radiative forcing targets of the representative concentration pathways in a world with agricultural climate impacts. *Earth's Futur.*, **2**, 83–98, doi:10.1002/2013EF000199. <http://doi.wiley.com/10.1002/2013EF000199>.
- Labrière, N., B. Locatelli, Y. Laumonier, V. Freycon, and M. Bernoux, 2015: Soil erosion in the humid tropics: A systematic quantitative review. *Agric. Ecosyst. Environ.*, **203**, 127–139, doi:10.1016/j.agee.2015.01.027. <https://linkinghub.elsevier.com/retrieve/pii/S0167880915000468> (Accessed November 10, 2018).
- Lal, R., 1998: Soil Erosion Impact on Agronomic Productivity and Environment Quality. *CRC. Crit. Rev. Plant Sci.*, **17**, 319–464, doi:10.1080/07352689891304249. <https://www.tandfonline.com/doi/full/10.1080/07352689891304249> (Accessed November 12, 2018).
- , 2001a: Soil degradation by erosion. *L. Degrad. Dev.*, **12**, 519–539, doi:10.1002/ldr.472. <http://doi.wiley.com/10.1002/ldr.472> (Accessed June 5, 2018).
- , 2001b: Soil degradation by erosion. *L. Degrad. Dev.*, **12**, 519–539, doi:10.1002/ldr.472.
- Lal, R. 2003. Soil erosion and the global carbon budget. *Environment International* 29 (2003) 437– 450.
- , 2006a: Enhancing crop yields in the developing countries through restoration of the soil organic carbon pool in agricultural lands. *L. Degrad. Dev.*, **17**, 197–209, doi:10.1002/ldr.696. <http://doi.wiley.com/10.1002/ldr.696> (Accessed June 5, 2018).
- , 2006b: Enhancing crop yields in the developing countries through restoration of the soil organic carbon pool in agricultural lands. *L. Degrad. Dev.*, **17**, 197–209, doi:10.1002/ldr.696. <http://doi.wiley.com/10.1002/ldr.696> (Accessed April 16, 2019).
- Lal, R., 2006c: Enhancing crop yields in the developing countries through restoration of the soil organic carbon pool in agricultural lands. *L. Degrad. Dev.*, **17**, 197–209, doi:10.1002/ldr.696. <http://doi.wiley.com/10.1002/ldr.696> (Accessed June 5, 2018).
- Lal, R., 2011: Sequestering carbon in soils of agro-ecosystems. *Food Policy*, **36**, S33–S39. <https://www.sciencedirect.com/science/article/pii/S0306919210001454> (Accessed November 11, 2018).
- Lal, R., 2014: Soil Carbon Management and Climate Change. *Soil Carbon*.

- Lal, R., 2015: Restoring Soil Quality to Mitigate Soil Degradation. *Sustainability*, **7**, 5875–5895, doi:10.3390/su7055875. <http://www.mdpi.com/2071-1050/7/5/5875> (Accessed June 5, 2018).
- Lal, R. Managing Soils and Ecosystems for Mitigating Anthropogenic Carbon Emissions and Advancing Global Food Security. *Bioscience* (2010).
- Lal, R., 2016: Soil health and carbon management. *Food Energy Secur.*, **5**, 212–222, doi:10.1002/fes3.96. <http://doi.wiley.com/10.1002/fes3.96> (Accessed November 10, 2018).
- , and W. C. Moldenhauer, 1987: Effects of soil erosion on crop productivity. *CRC. Crit. Rev. Plant Sci.*, **5**, 303–367, doi:10.1080/07352688709382244. <http://www.tandfonline.com/doi/abs/10.1080/07352688709382244> (Accessed November 11, 2018).
- Lamb, A., and Coauthors, 2016: The potential for land sparing to offset greenhouse gas emissions from agriculture. *Nat. Clim. Chang.*, **6**, 488–492, doi:10.1038/nclimate2910. <http://www.nature.com/articles/nclimate2910> (Accessed June 6, 2018).
- Lambin, E. F., and P. Meyfroidt, 2011: Global land use change, economic globalization, and the looming land scarcity. *Proc. Natl. Acad. Sci. U. S. A.*, **108**, 3465–3472, doi:10.1073/pnas.1100480108. <http://www.ncbi.nlm.nih.gov/pubmed/21321211> (Accessed May 30, 2018).
- Lambrou, Y., and G. Piana, 2006: Gender: The Missing Component of the Response to Climate Change. *Food Agric. Organ. United Nations*,.
- Lamers, P., E. Searcy, J. R. Hess, and H. Stichnothe, *Developing the global bioeconomy: technical, market, and environmental lessons from bioenergy*. 197 pp.
- Lamontagne, J. R., P. M. Reed, R. Link, K. V. Calvin, L. E. Clarke, and J. A. Edmonds, 2018: Large Ensemble Analytic Framework for Consequence-Driven Discovery of Climate Change Scenarios. *Earth's Futur.*, **6**, 488–504, doi:10.1002/2017EF000701. <http://doi.wiley.com/10.1002/2017EF000701> (Accessed April 14, 2019).
- Langford, W. T., A. Gordon, L. Bastin, S. A. Bekessy, M. D. White, and G. Newell, 2011: Raising the bar for systematic conservation planning. *Trends Ecol. Evol.*, **26**, 634–640, doi:10.1016/J.TREE.2011.08.001. <https://www.sciencedirect.com/science/article/pii/S0169534711002333> (Accessed April 4, 2019).
- Larsen, F. W., W. R. Turner, and T. M. Brooks, 2012: Conserving Critical Sites for Biodiversity Provides Disproportionate Benefits to People. *PLoS One*, **7**, e36971, doi:10.1371/journal.pone.0036971. <http://dx.plos.org/10.1371/journal.pone.0036971> (Accessed April 4, 2019).
- Larsen, S., N. S. Bentsen, T. Dalgaard, U. Jørgensen, J. E. Olesen, and C. Felby, 2017: Possibilities for near-term bioenergy production and GHG-mitigation through sustainable intensification of agriculture and forestry in Denmark. *Environ. Res. Lett.*, **12**, 114032.
- Lasco, R. D., R. J. P. Delfino, D. C. Catacutan, E. S. Simelton, and D. M. Wilson, 2014: Climate risk adaptation by smallholder farmers: The roles of trees and agroforestry. *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2013.11.013.
- Lauterjung, J., U. Münch, and A. Rudloff, 2010: The challenge of installing a tsunami early warning system in the vicinity of the Sunda Arc, Indonesia. *Hazards Earth Syst. Sci.*, **10**, 641–646, doi:10.5194/nhess-10-641-2010. www.nat-hazards-earth-syst-sci.net/10/641/2010/ (Accessed April 16, 2019).
- Lee-Smith, D., 2010: Cities feeding people: An update on urban agriculture in equatorial Africa. *Environ. Urban.*, doi:10.1177/0956247810377383.
- Lee, J., 2017: Farmer participation in a climate-smart future: Evidence from the Kenya Agricultural

- Carbon Project. *Land use policy*, **68**, 72–79, doi:10.1016/j.landusepol.2017.07.020. <https://linkinghub.elsevier.com/retrieve/pii/S0264837716307918> (Accessed November 10, 2018).
- Lee, J. H. W., I. J. Hodgkiss, K. T. M. Wong, and I. H. Y. Lam, 2005: Real time observations of coastal algal blooms by an early warning system. *Estuar. Coast. Shelf Sci.*, **65**, 172–190, doi:10.1016/J.ECSS.2005.06.005. <https://www.sciencedirect.com/science/article/pii/S0272771405001757> (Accessed April 16, 2019).
- Lee, Y., J. Ahern, and C. Yeh, 2015: Ecosystem services in peri-urban landscapes: The effects of agricultural landscape change on ecosystem services in Taiwan's western coastal plain. *Landsc. Urban Plan.*, <https://www.sciencedirect.com/science/article/pii/S0169204615000596> (Accessed May 1, 2018).
- Lejeune, Q., E. Davin, ... L. G.-N. C., and undefined 2018, Historical deforestation locally increased the intensity of hot days in northern mid-latitudes. *nature.com.*, <https://www.nature.com/articles/s41558-018-0131-z> (Accessed November 10, 2018).
- Lele, S., O. Springate-Baginski, R. Lakerveld, D. Deb, and P. Dash, 2013: Ecosystem Services: Origins, Contributions, Pitfalls, and Alternatives. *Conserv. Soc.*, **11**, 343, doi:10.4103/0972-4923.125752. <http://www.conservationandsociety.org/text.asp?2013/11/4/343/125752> (Accessed April 5, 2019).
- Lenton, T. M., 2010: The potential for land-based biological CO₂ removal to lower future atmospheric CO₂ concentration. *Carbon Manag.*, **1**, 145–160, doi:10.4155/cmt.10.12. <http://www.tandfonline.com/doi/abs/10.4155/cmt.10.12> (Accessed June 6, 2018).
- Lenton, T. The Global Potential for Carbon Dioxide Removal. in *Geoengineering of the Climate System* (eds. Harrison, R. M. & Hester, R. E.) 52–79 (Royal Society of Chemistry, 2014).
- Leskinen, P., and Coauthors, 2018: *Substitution effects of wood-based products in climate change mitigation.* https://www.alphagalileo.org/Uploads/documents/b4e54279-5b61-4b5f-a303-099de994ad0f-efi_fstp_7_2018.pdf (Accessed April 5, 2019).
- Lestrelin, G., and M. Giordano, 2007: Upland development policy, livelihood change and land degradation: interactions from a Laotian village. *L. Degrad. Dev.*, **18**, 55–76, doi:10.1002/ldr.756. <http://doi.wiley.com/10.1002/ldr.756> (Accessed November 11, 2018).
- Lewis, K., and C. Witham, 2012: Agricultural commodities and climate change. *Clim. Policy*, **12**, S53–S61, doi:10.1080/14693062.2012.728790. <http://www.tandfonline.com/doi/abs/10.1080/14693062.2012.728790> (Accessed June 1, 2018).
- Lewis, S. L., D. P. Edwards, and D. Galbraith, 2015: Increasing human dominance of tropical forests. *Science (80-.)*, doi:10.1126/science.aaa9932.
- , C. E. Wheeler, E. T. A. Mitchard, and A. Koch, 2019: Restoring natural forests is the best way to remove atmospheric carbon. *Nature*, **568**, 25–28, doi:10.1038/d41586-019-01026-8. <http://www.nature.com/articles/d41586-019-01026-8> (Accessed April 13, 2019).
- Li, J., R. Xu, D. Tiwari, and G. Ji, 2006: Effect of low-molecular-weight organic acids on the distribution of mobilized Al between soil solution and solid phase. *Appl. Geochemistry*, **21**, 1750–1759, doi:10.1016/J.APGEOCHEM.2006.06.013. <https://www.sciencedirect.com/science/article/pii/S0883292706001806> (Accessed April 12, 2019).
- Li, Y., M. Zhao, S. Motesharrei, Q. Mu, E. Kalnay, and S. Li, 2015: Local cooling and warming effects of forests based on satellite observations. *Nat. Commun.*, doi:10.1038/ncomms7603.
- Lillesø, J. B. L., and Coauthors, 2011: Innovation in input supply systems in smallholder agroforestry: Seed sources, supply chains and support systems. *Agrofor. Syst.*, doi:10.1007/s10457-011-9412-5.

- Limpens, J., and Coauthors, 2008: Peatlands and the carbon cycle: from local processes to global implications – a synthesis. *Biogeosciences*, **5**, 1475–1491. www.biogeosciences.net/5/1475/2008/ (Accessed June 7, 2018).
- Lin, B., Resilience in agriculture through crop diversification: adaptive management for environmental change. *academic.oup.com*,. <https://academic.oup.com/bioscience/article-abstract/61/3/183/238071> (Accessed November 12, 2018).
- Lin, B. B., 2011: Resilience in Agriculture through Crop Diversification: Adaptive Management for Environmental Change. *Bioscience*, **61**, 183–193, doi:10.1525/bio.2011.61.3.4. <https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2011.61.3.4> (Accessed April 16, 2019).
- Lin, C. S. K., and Coauthors, 2013: Food waste as a valuable resource for the production of chemicals, materials and fuels. Current situation and global perspective. *Energy Environ. Sci.*, **6**, 426, doi:10.1039/c2ee23440h. <http://xlink.rsc.org/?DOI=c2ee23440h> (Accessed April 16, 2019).
- Lin, Y., L. S. Wijedasa, and R. A. Chisholm, 2017: Singapore’s willingness to pay for mitigation of transboundary forest-fire haze from Indonesia. *Environ. Res. Lett.*, **12**, 24017, doi:10.1088/1748-9326/aa5cf6. <http://stacks.iop.org/1748-9326/12/i=2/a=024017?key=crossref.9c367504ef24d9670e82815ec04e5dfc> (Accessed May 31, 2018).
- Lipper, L., and Coauthors, 2014: Climate-smart agriculture for food security. *Nat. Clim. Chang.*, doi:10.1038/nclimate2437.
- Little, P. D., K. Smith, B. A. Cellarius, D. L. Coppock, and C. Barrett, 2001: Avoiding Disaster: Diversification and Risk Management among East African Herders. *Dev. Change*, **32**, 401–433, doi:10.1111/1467-7660.00211. <http://doi.wiley.com/10.1111/1467-7660.00211> (Accessed April 19, 2019).
- Liu, J., J. Lundqvist, J. Weinberg, and J. Gustafsson, 2013: Food Losses and Waste in China and Their Implication for Water and Land. *Environ. Sci. Technol.*, **47**, 10137–10144, doi:10.1021/es401426b. <http://pubs.acs.org/doi/abs/10.1021/es401426b> (Accessed April 16, 2019).
- , and Coauthors, 2017: Water scarcity assessments in the past, present, and future. *Earth’s Futur.*, **5**, 545–559, doi:10.1002/2016EF000518. <https://doi.org/10.1002/2016EF000518>.
- Liu, Y., Y. Zhou, and W. Wu, 2015: Assessing the impact of population, income and technology on energy consumption and industrial pollutant emissions in China. *Appl. Energy*, **155**, 904–917, doi:10.1016/J.APENERGY.2015.06.051. <https://www.sciencedirect.com/science/article/pii/S0306261915008090> (Accessed May 31, 2018).
- Liu, Z., and J. Lan, 2015: The Sloping Land Conversion Program in China: Effect on the Livelihood Diversification of Rural Households. *World Dev.*, **70**, 147–161, doi:10.1016/j.worlddev.2015.01.004. <https://linkinghub.elsevier.com/retrieve/pii/S0305750X15000054> (Accessed November 11, 2018).
- Lobell, D., M. Burke, ... C. T., and undefined 2008, Prioritizing climate change adaptation needs for food security in 2030. *science.sciencemag.org*, <http://science.sciencemag.org/content/319/5863/607.short> (Accessed November 11, 2018).
- Lobell, D. B., 2014: Climate change adaptation in crop production: Beware of illusions. *Glob. Food Sec.*, **3**, 72–76, doi:10.1016/j.gfs.2014.05.002. <https://www.sciencedirect.com/science/article/pii/S2211912414000145> (Accessed November 8, 2018).
- Lobell, D. B., U. L. C. Baldos, and T. W. Hertel, 2013: Climate adaptation as mitigation: the case of

- agricultural investments. *Environ. Res. Lett.*, **8**, 15012, doi:10.1088/1748-9326/8/1/015012. <http://stacks.iop.org/1748-9326/8/i=1/a=015012?key=crossref.75c10c4ca86540629b7564af461cd71f> (Accessed April 4, 2019).
- Locatelli, B., 2011: Synergies between adaptation and mitigation in a nutshell. *Cobam*, **8**.
- , V. Evans, A. Wardell, A. Andrade, and R. Vignola, 2011: Forests and climate change in latin America: Linking adaptation and mitigation. *Forests*, **2**, 431–450, doi:10.3390/f2010431.
- , and Coauthors, 2015a: Tropical reforestation and climate change: beyond carbon. *Restor. Ecol.*, **23**, 337–343, doi:10.1111/rec.12209. <http://doi.wiley.com/10.1111/rec.12209> (Accessed April 4, 2019).
- , and Coauthors, 2015b: Tropical reforestation and climate change: Beyond carbon. *Restor. Ecol.*, doi:10.1111/rec.12209.
- Locatelli, B., C. Pavageau, E. Pramova, and M. Di Gregorio, 2015c: Integrating climate change mitigation and adaptation in agriculture and forestry: Opportunities and trade-offs. *Wiley Interdiscip. Rev. Clim. Chang.*, **6**, 585–598, doi:10.1002/wcc.357.
- Locatelli, B., C. Pavageau, E. Pramova, and M. Di Gregorio, 2015d: Integrating climate change mitigation and adaptation in agriculture and forestry: opportunities and trade-offs. *Wiley Interdiscip. Rev. Clim. Chang.*, **6**, 585–598, doi:10.1002/wcc.357. <http://doi.wiley.com/10.1002/wcc.357> (Accessed April 16, 2019).
- Lock, K., J. Pomerleau, L. Causer, D. R. Altmann, and M. McKee, 2005: The global burden of disease attributable to low consumption of fruit and vegetables: implications for the global strategy on diet. *Bull. World Health Organ.*, doi:/S0042-96862005000200010.
- Logan, J. A., J. Régnière, and J. A. Powell, 2003: Assessing the impacts of global warming on forest pest dynamics. *Front. Ecol. Environ.*, **1**, 130–137, doi:10.1890/1540-9295(2003)001[0130:ATIOWG]2.0.CO;2. <https://esajournals.onlinelibrary.wiley.com/doi/full/10.1890/1540-9295%282003%29001%5B0130%3AATIOWG%5D2.0.CO%3B2> (Accessed April 4, 2019).
- Lopez, R., 2004: Urban Sprawl and Risk for Being Overweight or Obese. *Am. J. Public Health*, **94**, 1574–1579, doi:10.2105/AJPH.94.9.1574. <http://www.ncbi.nlm.nih.gov/pubmed/15333317> (Accessed April 16, 2019).
- Lopoukhine, N., and Coauthors, 2012: Protected areas: providing natural solutions to 21st Century challenges. <http://sapiens.revues.org/1254>.
- Lotze-Campen, H., and Coauthors, 2013: Impacts of increased bioenergy demand on global food markets: an AgMIP economic model intercomparison. *Agric. Econ.*, **45**, 103–116, doi:10.1111/agec.12092. <https://doi.org/10.1111/agec.12092>.
- Lotze, H., and Coauthors, 2006: Depletion, Degradation, and Recovery Potential of Estuaries and Coastal Seas. *Science (80-)*, **312**, 1806–1809, doi:10.1126/science.1128035. <https://www.sciencedirect.com/science/article/pii/S0028393206000868> (Accessed June 7, 2018).
- Louwaars, N. P., 2002: Seed Policy, Legislation and Law. *J. New Seeds*, **4**, 1–14, doi:10.1300/J153v04n01_01. http://www.tandfonline.com/doi/abs/10.1300/J153v04n01_01 (Accessed June 1, 2018).
- Lowder, S. K., J. Skoet, and T. Raney, 2016: The Number, Size, and Distribution of Farms, Smallholder Farms, and Family Farms Worldwide. *World Dev.*, **87**, 16–29, doi:10.1016/j.worlddev.2015.10.041. <https://linkinghub.elsevier.com/retrieve/pii/S0305750X15002703> (Accessed November 11, 2018).
- Lu, C., T. Zhao, X. Shi, and S. Cao, 2016: Ecological restoration by afforestation may increase

- groundwater depth and create potentially large ecological and water opportunity costs in arid and semiarid China. *J. Clean. Prod.*, **176**, 1213–1222, doi:<https://doi.org/10.1016/j.jclepro.2016.03.046>. <https://www.sciencedirect.com/science/article/pii/S0959652616301147> (Accessed June 7, 2018).
- Luby, C. H., J. Kloppenburg, T. E. Michaels, and I. L. Goldman, 2015: Enhancing Freedom to Operate for Plant Breeders and Farmers through Open Source Plant Breeding. *Crop Sci.*, **55**, 2481, doi:10.2135/cropsci2014.10.0708. <https://dl.sciencesocieties.org/publications/cs/abstracts/55/6/2481> (Accessed June 1, 2018).
- Luedeling, E., R. Kindt, N. I. Huth, and K. Koenig, 2014: Agroforestry systems in a changing climate—challenges in projecting future performance. *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2013.07.013.
- Lugato, E., K. Paustian, P. Panagos, A. Jones, and P. Borrelli, 2016: Quantifying the erosion effect on current carbon budget of European agricultural soils at high spatial resolution. *Glob. Chang. Biol.*, **22**, 1976–1984, doi:10.1111/gcb.13198. <http://doi.wiley.com/10.1111/gcb.13198> (Accessed June 5, 2018).
- Lundmark, T., and Coauthors, 2014: Potential Roles of Swedish Forestry in the Context of Climate Change Mitigation. *For. 2014, Vol. 5, Pages 557-578*, **5**, 557–578, doi:10.3390/F5040557.
- Lundy, M., C. F. O. Gálvez, and R. Best, 2002: *Value adding, agroenterprise and poverty reduction: A territorial approach for rural business development*.
- Luo, L., Y. Wang, and L. Qin, 2014: Incentives for promoting agricultural clean production technologies in China. *J. Clean. Prod.*, **74**, 54–61, doi:10.1016/J.JCLEPRO.2014.03.045. <https://www.sciencedirect.com/science/article/pii/S0959652614002662> (Accessed April 16, 2019).
- Luyssaert, S., and Coauthors, 2018: Trade-offs in using European forests to meet climate objectives. *Nature*, **562**, 259–262, doi:10.1038/s41586-018-0577-1. <http://www.nature.com/articles/s41586-018-0577-1> (Accessed April 5, 2019).
- Lwasa, S., F. Mugagga, B. Wahab, D. Simon, and J. C. Climate, 2014: Urban and peri-urban agriculture and forestry: transcending poverty alleviation to climate change mitigation and adaptation. *Urban Clim.*, <https://www.sciencedirect.com/science/article/pii/S2212095513000552> (Accessed May 1, 2018).
- , ———, ———, and ... D. S., 2015: A meta-analysis of urban and peri-urban agriculture and forestry in mediating climate change. *Curr. Opin. Environ. Sustain.*, <https://www.sciencedirect.com/science/article/pii/S1877343515000160> (Accessed May 1, 2018).
- Maaroufi, N. I., A. Nordin, N. J. Hasselquist, L. H. Bach, K. Palmqvist, and M. J. Gundale, 2015: Anthropogenic nitrogen deposition enhances carbon sequestration in boreal soils. *Glob. Chang. Biol.*, **21**, 3169–3180, doi:10.1111/gcb.12904. <http://doi.wiley.com/10.1111/gcb.12904> (Accessed November 11, 2018).
- Macdiarmid, J. I., H. Clark, S. Whybrow, H. de Ruiter, and G. McNeill, 2018: Assessing national nutrition security: The UK reliance on imports to meet population energy and nutrient recommendations. *PLoS One*, **13**, e0192649, doi:10.1371/journal.pone.0192649. <http://dx.plos.org/10.1371/journal.pone.0192649> (Accessed May 30, 2018).
- MacGregor, J., and B. Vorley, 2006: *Fair Miles? The concept of “food miles” through a sustainable development lens*.
- Machado, R., and R. Serralheiro, 2017a: Soil Salinity: Effect on Vegetable Crop Growth. Management Practices to Prevent and Mitigate Soil Salinization. *Horticulturae*, **3**, 30, doi:10.3390/horticulturae3020030. <http://www.mdpi.com/2311-7524/3/2/30>.

- , and ——, 2017b: Soil Salinity: Effect on Vegetable Crop Growth. Management Practices to Prevent and Mitigate Soil Salinization. *Horticulturae*, **3**, 30, doi:10.3390/horticulturae3020030. <http://www.mdpi.com/2311-7524/3/2/30> (Accessed April 18, 2019).
- Madlener, R., C. Robledo, B. Muys, and J. T. B. Freja, 2006: A Sustainability Framework for Enhancing the Long-Term Success of Lulucf Projects. *Clim. Change*, **75**, 241–271, doi:10.1007/s10584-005-9023-0. <http://link.springer.com/10.1007/s10584-005-9023-0> (Accessed June 7, 2018).
- Maes, J., and Coauthors, 2017a: Landslide risk reduction measures: A review of practices and challenges for the tropics. *Prog. Phys. Geogr.*, **41**, 191–221, doi:10.1177/0309133316689344. <http://journals.sagepub.com/doi/10.1177/0309133316689344> (Accessed May 31, 2018).
- , and Coauthors, 2017b: Landslide risk reduction measures: A review of practices and challenges for the tropics. *Prog. Phys. Geogr. Earth Environ.*, **41**, 191–221, doi:10.1177/0309133316689344. <http://journals.sagepub.com/doi/10.1177/0309133316689344> (Accessed April 18, 2019).
- Mahmood, R., and Coauthors, 2014: Land cover changes and their biogeophysical effects on climate. *Int. J. Climatol.*, doi:10.1002/joc.3736.
- Mahmud, T., and M. Prowse, 2012: Corruption in cyclone preparedness and relief efforts in coastal Bangladesh: Lessons for climate adaptation? *Glob. Environ. Chang.*, **22**, 933–943, doi:10.1016/J.GLOENVCHA.2012.07.003. <https://www.sciencedirect.com/science/article/pii/S0959378012000799> (Accessed June 1, 2018).
- Mal, S., R. B. Singh, C. Huggel, and A. Grover, 2018: Introducing Linkages Between Climate Change, Extreme Events, and Disaster Risk Reduction. Springer, Cham, 1–14 http://link.springer.com/10.1007/978-3-319-56469-2_1 (Accessed May 31, 2018).
- Maltsoglou, I., and Coauthors, 2014: Combining bioenergy and food security: An approach and rapid appraisal to guide bioenergy policy formulation. *Biomass and Bioenergy*, doi:10.1016/j.biombioe.2015.02.007.
- Manning, P., G. Taylor, and M. E. Hanley, 2015: Bioenergy, Food Production and Biodiversity - An Unlikely Alliance? *GCB Bioenergy*, **7**, 570–576, doi:10.1111/gcbb.12173.
- Margono, B. A., P. V. Potapov, S. Turubanova, F. Stolle, and M. C. Hansen, 2014: Primary forest cover loss in Indonesia over 2000–2012. *Nat. Clim. Chang.*, **4**, 730–735, doi:10.1038/nclimate2277. <http://www.nature.com/doi/10.1038/nclimate2277> (Accessed June 2, 2018).
- Mariola, M. J., 2008: The local industrial complex? Questioning the link between local foods and energy use. *Agric. Human Values*, **25**, 193–196, doi:10.1007/s10460-008-9115-3. <http://link.springer.com/10.1007/s10460-008-9115-3> (Accessed April 16, 2019).
- Markandya, A., J. Sampedro, S. J. Smith, R. Van Dingenen, C. Pizarro-Irizar, I. Arto, and M. González-Eguino, 2018: Health co-benefits from air pollution and mitigation costs of the Paris Agreement: a modelling study. *Lancet Planet. Heal.*, **2**, e126–e133, doi:10.1016/S2542-5196(18)30029-9. <https://www.sciencedirect.com/science/article/pii/S2542519618300299> (Accessed June 1, 2018).
- Marques, M., and Coauthors, 2016: Multifaceted Impacts of Sustainable Land Management in Drylands: A Review. *Sustainability*, **8**, 177, doi:10.3390/su8020177. <http://www.mdpi.com/2071-1050/8/2/177> (Accessed November 10, 2018).
- Martha, G. B., E. Alves, and E. Contini, 2012: Land-saving approaches and beef production growth in Brazil. *Agric. Syst.*, **110**, 173–177, doi:10.1016/J.AGSY.2012.03.001. <https://www.sciencedirect.com/science/article/pii/S0308521X12000340> (Accessed June 7, 2018).
- Martin, A., N. Gross-Camp, and A. Akol, 2015: Towards an Explicit Justice Framing of the Social Impacts of Conservation. *Conserv. Soc.*, **13**, 166, doi:10.4103/0972-4923.164200.

- <http://www.conservationandsociety.org/text.asp?2015/13/2/166/164200> (Accessed June 7, 2018).
- Martin, S. M., and K. Lorenzen, 2016: Livelihood Diversification in Rural Laos. *World Dev.*, **83**, 231–243, doi:10.1016/J.WORLDDEV.2016.01.018. <https://www.sciencedirect.com/science/article/pii/S0305750X16000164> (Accessed May 31, 2018).
- Maskrey, A., 2011: Revisiting community-based disaster risk management. *Environ. Hazards*, **10**, 42–52, doi:10.3763/ehaz.2011.0005.
- Massawe, F., S. Mayes, and A. Cheng, 2016: Crop Diversity: An Unexploited Treasure Trove for Food Security. *Trends Plant Sci.*, **21**, 365–368, doi:10.1016/J.TPLANTS.2016.02.006. <https://www.sciencedirect.com/science/article/pii/S1360138516000601> (Accessed November 9, 2018).
- Mathbor, G. M., 2007: Enhancement of community preparedness for natural disasters. *Int. Soc. Work*, **50**, 357–369, doi:10.1177/0020872807076049. <http://journals.sagepub.com/doi/10.1177/0020872807076049> (Accessed April 16, 2019).
- Mathijs, E., 2015: Exploring future patterns of meat consumption. *Meat Sci.*, **109**, 112–116, doi:10.1016/J.MEATSCI.2015.05.007. <https://www.sciencedirect.com/science/article/abs/pii/S030917401530005X> (Accessed April 16, 2019).
- Maxwell, D., 1999: The Political Economy of Urban Food Security in Sub-Saharan Africa. *World Dev.*, **27**, 1939–1953, doi:10.1016/S0305-750X(99)00101-1. <https://www.sciencedirect.com/science/article/abs/pii/S0305750X99001011> (Accessed April 16, 2019).
- , and K. Wiebe, 1999: Land tenure and food security: Exploring dynamic linkages. *Dev. Change*, **30**, 825–849, doi:10.1111/1467-7660.00139. <http://doi.wiley.com/10.1111/1467-7660.00139> (Accessed June 7, 2018).
- Mazzeo, J., and B. Brenton, 2013: Peasant Resistance to Hybrid Seed in Haiti: The Implications of Humanitarian Aid on Food Security and Cultural Identity. *Food and Identity in the Caribbean*, London, 121–137.
- Mbow, C., M. Van Noordwijk, and P. A. Minang, 2014a: Agroforestry solutions to address food security and climate change challenges in Africa. *Curr. Opin. Environ. Sustain.*, **6**, 61–67, doi:10.1016/J.COSUST.2013.10.014. <https://www.sciencedirect.com/science/article/pii/S1877343513001449> (Accessed April 16, 2019).
- Mbow C., Smith P. Skole D., Duguma L., Bustamante M. 2014. Achieving mitigation and adaptation to climate change through sustainable agroforestry practices in Africa. *Current Opinion in Environmental Sustainability* 6:8–14. <http://dx.doi.org/10.1016/j.cosust.2013.09.002>
- , P. Smith, D. Skole, L. Duguma, and M. Bustamante, 2014b: Achieving mitigation and adaptation to climate change through sustainable agroforestry practices in Africa. *Curr. Opin. Environ. Sustain.*, **6**, 8–14, doi:10.1016/j.cosust.2013.09.002. <https://www.sciencedirect.com/science/article/pii/S1877343513001255> (Accessed May 30, 2018).
- , ———, ———, ———, and ———, 2014c: Achieving mitigation and adaptation to climate change through sustainable agroforestry practices in Africa. *Curr. Opin. Environ. Sustain.*, **6**, 8–14, doi:10.1016/j.cosust.2013.09.002. <https://www.sciencedirect.com/science/article/pii/S1877343513001255> (Accessed November 12, 2018).

- McCullum, D., N. Bauer, K. Calvin, A. Kitous, and K. Riahi, 2014: Fossil resource and energy security dynamics in conventional and carbon-constrained worlds. *Clim. Change*, **123**, doi:10.1007/s10584-013-0939-5.
- McCrum, G., K. Blackstock, K. Matthews, M. Rivington, D. Miller, and K. Buchan, 2009: Adapting to climate change in land management: the role of deliberative workshops in enhancing social learning. *Environ. Policy Gov.*, **19**, 413–426, doi:10.1002/eet.525. <http://doi.wiley.com/10.1002/eet.525> (Accessed December 26, 2017).
- McElwee, P., T. Nghiem, H. Le, and H. Vu, 2017a: Flood vulnerability among rural households in the Red River Delta of Vietnam: implications for future climate change risk and adaptation. *Nat. Hazards*, **86**, 465–492, doi:10.1007/s11069-016-2701-6. <http://link.springer.com/10.1007/s11069-016-2701-6> (Accessed June 7, 2018).
- , V. H. T. Nguyen, D. V. Nguyen, N. H. Tran, H. V. T. Le, T. P. Nghiem, and H. D. T. Vu, 2017b: Using REDD+ policy to facilitate climate adaptation at the local level: Synergies and challenges in Vietnam. *Forests*, **8**, 1–25, doi:10.3390/f8010011.
- McGuire, S., and L. Sperling, 2016: Seed systems smallholder farmers use. *Food Secur.*, **8**, 179–195, doi:10.1007/s12571-015-0528-8. <http://link.springer.com/10.1007/s12571-015-0528-8> (Accessed November 11, 2018).
- McKinsey and Company, 2009: *Pathways to a low-carbon economy: Version 2 of the global greenhouse gas abatement cost curve | McKinsey & Company*. <https://www.mckinsey.com/business-functions/sustainability-and-resource-productivity/our-insights/pathways-to-a-low-carbon-economy> (Accessed May 30, 2018).
- McLaren, D., 2012a: A comparative global assessment of potential negative emissions technologies. *Process Saf. Environ. Prot.*, **90**, 489–500, doi:10.1016/J.PSEP.2012.10.005. <https://www.sciencedirect.com/science/article/pii/S0957582012001176> (Accessed April 5, 2019).
- , 2012b: A comparative global assessment of potential negative emissions technologies. *Process Saf. Environ. Prot.*, **90**, 489–500, doi:10.1016/J.PSEP.2012.10.005. <https://www.sciencedirect.com/science/article/pii/S0957582012001176> (Accessed April 14, 2019).
- McLeman, R., and B. Smit, 2006: Migration as an adaptation to climate change. *Clim. Change*, **76**, 31–53. https://idp.springer.com/authorize/casa?redirect_uri=https://link.springer.com/article/10.1007/s10584-005-9000-7&casa_token=Uil_gsHqH98AAAAA:EEItzYZaPM8D-b3vh2ezMW5GP95nlVwk_ZRJUCDD5zEgXY30ikHi29YxddyqeCTTDQXxJnr-HBIReX2AL (Accessed November 11, 2018).
- McMichael, A. J., J. W. Powles, C. D. Butler, and R. Uauy, 2007: Food, livestock production, energy, climate change, and health. *Lancet*, **370**, 1253–1263, doi:10.1016/S0140-6736(07)61256-2. <https://www.sciencedirect.com/science/article/pii/S0140673607612562> (Accessed April 16, 2019).
- McMichael, P., 2012: The land grab and corporate food regime restructuring. *J. Peasant Stud.*, **39**, 681–701, doi:10.1080/03066150.2012.661369.
- , and M. Schneider, 2011a: Food Security Politics and the Millennium Development Goals. *Third World Q.*, **32**, 119–139, doi:10.1080/01436597.2011.543818.
- , and ———, 2011b: Food Security Politics and the Millennium Development Goals. *Third World Q.*, **32**, 119–139, doi:10.1080/01436597.2011.543818. <http://www.tandfonline.com/doi/abs/10.1080/01436597.2011.543818> (Accessed November 11, 2018).
- Mechler, R., 2016: Reviewing estimates of the economic efficiency of disaster risk management:

- opportunities and limitations of using risk-based cost–benefit analysis. *Nat. Hazards*, **81**, 2121–2147, doi:10.1007/s11069-016-2170-y. <http://link.springer.com/10.1007/s11069-016-2170-y> (Accessed April 18, 2019).
- Meier, P., D. Bond, and J. Bond, 2007: Environmental influences on pastoral conflict in the Horn of Africa. *Polit. Geogr.*, **26**, 716–735, doi:10.1016/J.POLGEO.2007.06.001. <https://www.sciencedirect.com/science/article/pii/S0962629807000820> (Accessed April 16, 2019).
- Meijer, S. S., D. Catacutan, O. C. Ajayi, G. W. Sileshi, and M. Nieuwenhuis, 2015a: The role of knowledge, attitudes and perceptions in the uptake of agricultural and agroforestry innovations among smallholder farmers in sub-Saharan Africa. *Int. J. Agric. Sustain.*, doi:10.1080/14735903.2014.912493.
- , ———, ———, ———, and ———, 2015b: The role of knowledge, attitudes and perceptions in the uptake of agricultural and agroforestry innovations among smallholder farmers in sub-Saharan Africa. *Int. J. Agric. Sustain.*, **13**, 40–54, doi:10.1080/14735903.2014.912493. <https://www.tandfonline.com/doi/full/10.1080/14735903.2014.912493> (Accessed April 18, 2019).
- van Meijl, H., and Coauthors, 2018: Comparing impacts of climate change and mitigation on global agriculture by 2050. *Environ. Res. Lett.*, **13**, 64021, doi:10.1088/1748-9326/aabdc4. <http://stacks.iop.org/1748-9326/13/i=6/a=064021?key=crossref.42a4eb1897f2ed545f2b0dc439d03e64> (Accessed April 14, 2019).
- Mekuria, W., and E. Aynekulu, 2013: EXCLOSURE LAND MANAGEMENT FOR RESTORATION OF THE SOILS IN DEGRADED COMMUNAL GRAZING LANDS IN NORTHERN ETHIOPIA. *L. Degrad. Dev.*, **24**, 528–538, doi:10.1002/ldr.1146. <http://doi.wiley.com/10.1002/ldr.1146> (Accessed November 12, 2018).
- Melamed, M., and J. Schmale, 2016: Sustainable policy—key considerations for air quality and climate change. *Curr. Opin. Environ. Sustain.*, **23**, 85–91. <https://www.sciencedirect.com/science/article/pii/S1877343516301087> (Accessed November 11, 2018).
- Mello, F. F. C., and Coauthors, 2014a: Payback time for soil carbon and sugar-cane ethanol. *Nat. Clim. Chang.*, **4**, 605–609, doi:10.1038/nclimate2239. <http://www.nature.com/articles/nclimate2239> (Accessed November 11, 2018).
- , and Coauthors, 2014b: Payback time for soil carbon and sugar-cane ethanol. *Nat. Clim. Chang.*, **4**, 605–609, doi:10.1038/nclimate2239.
- Mello, F. F. C., and Coauthors, 2014c: Payback time for soil carbon and sugar-cane ethanol. *Nat. Clim. Chang.*, doi:10.1038/nclimate2239.
- Melo, F. P. L., S. R. R. Pinto, P. H. S. Brancalion, P. S. Castro, R. R. Rodrigues, J. Aronson, and M. Tabarelli, 2013: Priority setting for scaling-up tropical forest restoration projects: Early lessons from the Atlantic Forest Restoration Pact. *Environ. Sci. Policy*, **33**, 395–404, doi:10.1016/j.envsci.2013.07.013.
- Mercer, J., 2010: Policy Arena Disaster Risk Reduction or Climate Change Adaptation: Are We Reinventing the Wheel? *J. Int. Dev.*, doi:10.1002/jid.
- Merriott, D., 2016: Factors associated with the farmer suicide crisis in India. *J. Epidemiol. Glob. Health*, **6**, 217–227, doi:10.1016/J.JEGH.2016.03.003. <https://www.sciencedirect.com/science/article/pii/S2210600615300277> (Accessed June 2, 2018).
- Meyer, S., B. Glaser, and P. Quicker, 2011: Technical, Economical, and Climate-Related Aspects of Biochar Production Technologies: A Literature Review. *Environ. Sci. Technol.*, **45**, 9473–9483,

- doi:10.1021/es201792c. <http://pubs.acs.org/doi/abs/10.1021/es201792c> (Accessed August 13, 2018).
- Meze-Hausken, E., A. Patt, and S. Fritz, 2009: Reducing climate risk for micro-insurance providers in Africa: A case study of Ethiopia. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2008.09.001.
- Miao, L., J. C. Moore, F. Zeng, J. Lei, J. Ding, B. He, and X. Cui, 2015: Footprint of Research in Desertification Management in China. *L. Degrad. Dev.*, **26**, 450–457, doi:10.1002/ldr.2399. <http://doi.wiley.com/10.1002/ldr.2399> (Accessed November 11, 2018).
- Michelini, L., L. Principato, and G. Iasevoli, 2018: Understanding Food Sharing Models to Tackle Sustainability Challenges. *Ecol. Econ.*, **145**, 205–217, doi:10.1016/J.ECOLECON.2017.09.009. <https://www.sciencedirect.com/science/article/pii/S0921800916308370> (Accessed June 1, 2018).
- Midmore, D. J., and H. G. P. Jansen, 2003: Supplying vegetables to Asian cities: is there a case for peri-urban production? *Food Policy*, **28**, 13–27, doi:10.1016/S0306-9192(02)00067-2. <https://www.sciencedirect.com/science/article/abs/pii/S0306919202000672> (Accessed April 16, 2019).
- Millenium Ecosystem Assessment (MA), 2005: *Ecosystems and Human Well-being: Synthesis*. Island Press, Washington, DC.,.
- Mills, E., 2005: Insurance in a climate of change. *Science*, **309**, 1040–1044, doi:10.1126/science.1112121. <http://www.ncbi.nlm.nih.gov/pubmed/16099975> (Accessed April 4, 2019).
- Miner, R., 2010: *Impact of the global forest industry on atmospheric greenhouse gases*. Food and Agriculture Organization of the United Nations (FAO),.
- Minot, N., 2014: Food price volatility in sub-Saharan Africa: Has it really increased? *Food Policy*, **45**, 45–56, doi:10.1016/J.FOODPOL.2013.12.008. <https://www.sciencedirect.com/science/article/pii/S0306919213001863> (Accessed June 1, 2018).
- Mittermeier, R. A., W. R. Turner, F. W. Larsen, T. M. Brooks, and C. Gascon, 2011: Global Biodiversity Conservation: The Critical Role of Hotspots. *Biodiversity Hotspots*, Springer Berlin Heidelberg, Berlin, Heidelberg, 3–22 http://link.springer.com/10.1007/978-3-642-20992-5_1 (Accessed April 13, 2019).
- Mohammadi, A., S. Rafiee, A. Jafari, A. Keyhani, S. H. Mousavi-Avval, and S. Nonhebel, 2014: Energy use efficiency and greenhouse gas emissions of farming systems in north Iran. *Renew. Sustain. Energy Rev.*, **30**, 724–733, doi:10.1016/J.RSER.2013.11.012. <https://www.sciencedirect.com/science/article/pii/S1364032113007612> (Accessed June 7, 2018).
- Molotoks, A., M. Kuhnert, T. Dawson, and P. Smith, 2017: Global Hotspots of Conflict Risk between Food Security and Biodiversity Conservation. *Land*, doi:10.3390/land6040067.
- Monier, E., and Coauthors, 2018: Toward a consistent modeling framework to assess multi-sectoral climate impacts. *Nat. Commun.*, **9**, 660, doi:10.1038/s41467-018-02984-9. <http://www.nature.com/articles/s41467-018-02984-9> (Accessed April 14, 2019).
- Monteiro, C. A., 2009: Nutrition and health. The issue is not food, nor nutrients, so much as processing. *Public Health Nutr.*, **12**, 729–731, doi:10.1017/S1368980009005291. https://www.cambridge.org/core/product/identifier/S1368980009005291/type/journal_article (Accessed April 16, 2019).
- Monteiro, C. A., R. B. Levy, R. M. Claro, I. R. R. de Castro, and G. Cannon, 2011: Increasing consumption of ultra-processed foods and likely impact on human health: evidence from Brazil. *Public Health Nutr.*, **14**, 5–13, doi:10.1017/S1368980010003241. http://www.journals.cambridge.org/abstract_S1368980010003241 (Accessed April 16, 2019).
- Mooney, H. A., and P. R. Ehrlich, 1997: Ecosystem services: A fragmentary history. *Nature's services* :

- societal dependence on natural ecosystems*, Island Press, p. 392 <https://books.google.co.nz/books?hl=en&lr=&id=QYJSziDfTjEC&oi=fnd&pg=PA11&dq=.+Ecosystem+services:+A+fragmentary+history&ots=YgyMNQIYDm&sig=bxhNCacCZP5xrjaKa4sK00dgMT4#v=onepage&q=.+Ecosystem+services%3A+A+fragmentary+history&f=false> (Accessed April 5, 2019).
- Moore, D. J., J. E. Tilton, and D. J. Shields, 1996: Economic growth and the demand for construction materials. *Resour. Policy*, **22**, 197–205, doi:10.1016/S0301-4207(96)00037-2. <https://www.sciencedirect.com/science/article/abs/pii/S0301420796000372> (Accessed April 16, 2019).
- de Moraes Sá, J. C., R. Lal, C. C. Cerri, K. Lorenz, M. Hungria, and P. C. de Faccio Carvalho, 2017: Low-carbon agriculture in South America to mitigate global climate change and advance food security. *Environ. Int.*, **98**, 102–112, doi:10.1016/J.ENVINT.2016.10.020. <https://www.sciencedirect.com/science/article/pii/S0160412016306341> (Accessed May 31, 2018).
- Morduch, J., and M. Sharma, 2002: Strengthening Public Safety Nets from the Bottom Up. *Dev. Policy Rev.*, doi:10.1111/1467-7679.00190.
- Morton, J. F., 2007: The impact of climate change on smallholder and subsistence agriculture. *Proc. Natl. Acad. Sci. U. S. A.*, **104**, 19680–19685, doi:10.1073/pnas.0701855104.
- Mosquera-Losada, M. R.; Santiago-Freijanes, J. J.; Pisanelli, A.; Rois-Díaz, M.; Smith, J.; den Herder, M.; Moreno, G.; Ferreira-Domínguez, N.; Malignier, N.; Lamersdorf, N.; Balaguer, F.; Pantera, A.; Rigueiro-Rodríguez, A.; Aldrey, J. A.; González-Hernández, M. P.; Fernández-Lorenzo, J. L.; Romero-Franco, R. & Burgess, P. J. 2018. Agroforestry in the European common agricultural policy. *Agroforest Syst* (2018) 92: 1117. <https://doi.org/10.1007/s10457-018-0251-5>
- Moss, R. H., A. L. Brenkert, and E. L. Malone, 2001: Vulnerability to Climate Change. A Quantitative Approach. *Prep. US Dep. Energy Available online <http://www.glob.umd.edu/cgibin/Details.pl>,*
- Mostofa, K., C. Liu, W. Zhai, M. Minella, D. V.- Biogeosciences, and undefined 2016, Reviews and Syntheses: Ocean acidification and its potential impacts on marine ecosystems. *iris.unito.it*, <https://iris.unito.it/handle/2318/1616088> (Accessed November 11, 2018).
- Mouratiadou, I., and Coauthors, 2016: The impact of climate change mitigation on water demand for energy and food: An integrated analysis based on the Shared Socioeconomic Pathways. *Environ. Sci. Policy*, **64**, 48–58, doi:10.1016/j.envsci.2016.06.007. <http://dx.doi.org/10.1016/j.envsci.2016.06.007>.
- Mousseau, F., 2015: The Untold Success Story of Agroecology in Africa. *Development*, **58**, 341–345, doi:10.1057/s41301-016-0026-0. <http://link.springer.com/10.1057/s41301-016-0026-0> (Accessed November 11, 2018).
- Mudombi, S., and Coauthors, 2018: Multi-dimensional poverty effects around operational biofuel projects in Malawi, Mozambique and Swaziland. *Biomass and bioenergy*, **114**, 41–54.
- Mueller, N. D., J. S. Gerber, M. Johnston, D. K. Ray, N. Ramankutty, and J. A. Foley, 2012: Closing yield gaps through nutrient and water management. *Nature*, **490**, 254–257, doi:10.1038/nature11420. <http://www.nature.com/articles/nature11420> (Accessed April 16, 2019).
- Mujumdar, M., K. P. Sooraj, R. Krishnan, B. Preethi, M. K. Joshi, H. Varikoden, B. B. Singh, and M. Rajeevan, 2017: Anomalous convective activity over sub-tropical east Pacific during 2015 and associated boreal summer monsoon teleconnections. *Clim. Dyn.*, **48**, 4081–4091.
- Muller, A., and Coauthors, 2017a: Strategies for feeding the world more sustainably with organic

- agriculture. *Nat. Commun.*, **8**, 1290, doi:10.1038/s41467-017-01410-w. <http://www.nature.com/articles/s41467-017-01410-w> (Accessed November 8, 2018).
- , and Coauthors, 2017b: Strategies for feeding the world more sustainably with organic agriculture. *Nat. Commun.*, doi:10.1038/s41467-017-01410-w.
- Munang, R., J. Andrews, K. Alverson, and D. Mebratu, 2014: Harnessing ecosystem-based adaptation to address the social dimensions of climate change. *Environ. Sci. Policy Sustain. Dev.*, **56**, 18–24, doi:10.1080/00139157.2014.861676. <http://www.tandfonline.com/doi/abs/10.1080/00139157.2014.861676> (Accessed May 30, 2018).
- Mundler, P., and L. Rumpus, 2012: The energy efficiency of local food systems: A comparison between different modes of distribution. *Food Policy*, **37**, 609–615, doi:10.1016/J.FOODPOL.2012.07.006. <https://www.sciencedirect.com/science/article/pii/S0306919212000802> (Accessed June 1, 2018).
- Muratori, M., K. Calvin, M. Wise, P. Kyle, and J. Edmonds, 2016: Global economic consequences of deploying bioenergy with carbon capture and storage (BECCS). *Environ. Res. Lett.*, **11**, 95004, doi:10.1088/1748-9326/11/9/095004. <http://stacks.iop.org/1748-9326/11/i=9/a=095004?key=crossref.962cb328d4e5c48882928ffcb3cdc7fe>.
- Muriithi, B. W., and J. A. Matz, 2015: Welfare effects of vegetable commercialization: Evidence from smallholder producers in Kenya. *Food Policy*, **50**, 80–91, doi:10.1016/J.FOODPOL.2014.11.001. <https://www.sciencedirect.com/science/article/abs/pii/S0306919214001882> (Accessed April 16, 2019).
- Murthy, P. S., and M. Madhava Naidu, 2012: Sustainable management of coffee industry by-products and value addition - A review. *Resour. Conserv. Recycl.*, doi:10.1016/j.resconrec.2012.06.005.
- Mustafa, D., G. Gioli, S. Qazi, R. Waraich, A. Rehman, and R. Zahoor, 2015: Gendering flood early warning systems: the case of Pakistan. *Environ. Hazards*, **14**, 312–328, doi:10.1080/17477891.2015.1075859. <http://www.tandfonline.com/doi/full/10.1080/17477891.2015.1075859> (Accessed April 16, 2019).
- Mutoko, M., C. Shisanya, and L. Hein, 2014: Fostering technological transition to sustainable land management through stakeholder collaboration in the western highlands of Kenya. *Land use policy*, **41**, 110–120. <https://www.sciencedirect.com/science/article/pii/S0264837714001112> (Accessed November 10, 2018).
- Mutuo, P. K., G. Cadisch, A. Albrecht, C. A. Palm, and L. Verchot, 2005: Potential of agroforestry for carbon sequestration and mitigation of greenhouse gas emissions from soils in the tropics. *Nutr. Cycl. Agroecosystems*, **71**, 43–54, doi:10.1007/s10705-004-5285-6. <http://link.springer.com/10.1007/s10705-004-5285-6> (Accessed April 5, 2019).
- Nabuurs, G. J., A. Pussinen, J. van Brusselen, and M. J. Schelhaas, 2007: Future harvesting pressure on European forests. *Eur. J. For. Res.*, **126**, 391–400, doi:10.1007/s10342-006-0158-y. <http://link.springer.com/10.1007/s10342-006-0158-y> (Accessed April 13, 2019).
- Nabuurs, G. J., P. Delacote, D. Ellison, M. Hanewinkel, L. Hetemäki, M. Lindner, and M. Ollikainen, 2017: By 2050 the mitigation effects of EU forests could nearly double through climate smart forestry. *Forests*, **8**, 1–14, doi:10.3390/f8120484.
- Naeem, S., and Coauthors, 2015: Get the science right when paying for nature’s services. *Science (80-.)*, **347**, 1206–1207.
- Nagoda, S., and A. J. Nightingale, 2017: Participation and Power in Climate Change Adaptation Policies: Vulnerability in Food Security Programs in Nepal. *World Dev.*, doi:10.1016/j.worlddev.2017.07.022.

- Nahlik, A. M., M. E. Kentula, M. S. Fennessy, and D. H. Landers, 2012: Where is the consensus? A proposed foundation for moving ecosystem service concepts into practice. *Ecol. Econ.*, **77**, 27–35, doi:10.1016/J.ECOLECON.2012.01.001. <https://www.sciencedirect.com/science/article/pii/S092180091200002X> (Accessed April 5, 2019).
- Nahman, A., and W. de Lange, 2013: Costs of food waste along the value chain: Evidence from South Africa. *Waste Manag.*, **33**, 2493–2500, doi:10.1016/J.WASMAN.2013.07.012. <https://www.sciencedirect.com/science/article/pii/S0956053X13003401> (Accessed April 16, 2019).
- , ——, S. Oelofse, and L. Godfrey, 2012: The costs of household food waste in South Africa. *Waste Manag.*, **32**, 2147–2153, doi:10.1016/J.WASMAN.2012.04.012. <https://www.sciencedirect.com/science/article/pii/S0956053X12001766> (Accessed April 16, 2019).
- Nair, P. K. R., D. N. Vimala, M. S. Julia, and B. M. Kumar, 2010: Carbon Sequestration in Agroforestry Systems. *Adv. Agron.*, **108**, 237.
- Nandy, S., A. Daoud, and D. Gordon, 2016: Examining the changing profile of undernutrition in the context of food price rises and greater inequality. *Soc. Sci. Med.*, **149**, 153–163, doi:10.1016/j.socscimed.2015.11.036. <https://linkinghub.elsevier.com/retrieve/pii/S0277953615302446> (Accessed November 11, 2018).
- Narain, P., R. K. Singh, N. S. Sindhwal, and P. Joshie, 1997: Agroforestry for soil and water conservation in the western Himalayan Valley Region of India 2. Crop and tree production. *Agrofor. Syst.*, **39**, 191–203, doi:10.1023/A:1005900229886. <http://link.springer.com/10.1023/A:1005900229886> (Accessed June 7, 2018).
- Nardone, A., B. Ronchi, N. Lacetera, M. S. Ranieri, and U. Bernabucci, 2010: Effects of climate changes on animal production and sustainability of livestock systems. *Livest. Sci.*, **130**, 57–69, doi:10.1016/J.LIVSCI.2010.02.011. <https://www.sciencedirect.com/science/article/pii/S1871141310000740> (Accessed April 4, 2019).
- Naudts, K., Y. Chen, M. J. McGrath, J. Ryder, A. Valade, J. Otto, and S. Luyssaert, 2016: Europe’s forest management did not Mitigate Climate Warming. *Science (80-.)*, **351**, 597–601, doi:10.1126/science.aac9976.
- Naylor, R. L., and Coauthors, 2000a: Effect of aquaculture on world fish supplies. *Nature*, **405**, 1017–1024, doi:10.1038/35016500. <http://www.nature.com/articles/35016500> (Accessed May 30, 2018).
- , and Coauthors, 2000b: Effect of aquaculture on world fish supplies. *Nature*, **405**, 1017–1024, doi:10.1038/35016500. <http://www.nature.com/articles/35016500> (Accessed May 30, 2018).
- Neary, D. G., G. G. Ice, and C. R. Jackson, 2009a: Linkages between forest soils and water quality and quantity. *For. Ecol. Manage.*, **258**, 2269–2281, doi:10.1016/J.FORECO.2009.05.027. <https://www.sciencedirect.com/science/article/pii/S0378112709003557> (Accessed April 16, 2019).
- Neary, D. G., G. G. Ice, and C. R. Jackson, 2009b: Linkages between forest soils and water quality and quantity. *For. Ecol. Manage.*, doi:10.1016/j.foreco.2009.05.027. <https://www.sciencedirect.com/science/article/pii/S0378112709003557> (Accessed November 9, 2018).
- Neff, R. A., A. M. Palmer, S. E. McKenzie, and R. S. Lawrence, 2009: Food Systems and Public Health Disparities. *J. Hunger Environ. Nutr.*, **4**, 282–314, doi:10.1080/19320240903337041. <http://www.tandfonline.com/doi/abs/10.1080/19320240903337041> (Accessed April 16, 2019).
- Neilson, J., 2007: Institutions, the governance of quality and on-farm value retention for Indonesian specialty coffee. *Singap. J. Trop. Geogr.*, **28**, 188–204, doi:10.1111/j.1467-9493.2007.00290.x. <http://doi.wiley.com/10.1111/j.1467-9493.2007.00290.x> (Accessed April 16, 2019).

- Nejad, A. N., 2013: *Soil and water conservation for desertification control in Iran*. 377-400 pp.
- Nelson, G. C., and Coauthors, 2014: Climate change effects on agriculture: Economic responses to biophysical shocks. *Proc. Natl. Acad. Sci.*, **111**, 3274–3279, doi:10.1073/pnas.1222465110. <http://www.pnas.org/lookup/doi/10.1073/pnas.1222465110>.
- Nelson, V., K. Meadows, T. Cannon, J. Morton, and A. Martin, 2002: Uncertain predictions, invisible impacts, and the need to mainstream gender in climate change adaptations. *Gen. Dev.*, doi:10.1080/13552070215911.
- Nemet, G. F., T. Holloway, and P. Meier, 2010: Implications of incorporating air-quality co-benefits into climate change policymaking. *Environ. Res. Lett.*, **5**, 14007, doi:10.1088/1748-9326/5/1/014007. <http://stacks.iop.org/1748-9326/5/i=1/a=014007?key=crossref.2ba625d6bb5777ba4233ddfb8296f5eb> (Accessed November 11, 2018).
- Netzel, P., and T. Stepinski, 2018: Climate Similarity Search: GeoWeb Tool for Exploring Climate Variability. *Bull. Am. Meteorol. Soc.*, **99**, 475–477, doi:10.1175/BAMS-D-16-0334.1. <http://journals.ametsoc.org/doi/10.1175/BAMS-D-16-0334.1> (Accessed November 10, 2018).
- Newbold, T., L. Hudson, S. Hill, S. Contu, I. L.- Nature, and undefined 2015, Global effects of land use on local terrestrial biodiversity. *nature.com*,.
- Newbold, T., and Coauthors, 2015: Global effects of land use on local terrestrial biodiversity. *Nature*, **520**, 45–50, doi:10.1038/nature14324. <http://www.nature.com/articles/nature14324> (Accessed June 3, 2018).
- Newfarmer, R. S., W. Shaw, and P. Walkenhorst, 2009: *Breaking into new markets : emerging lessons for export diversification*. World Bank, 265 pp.
- Ngcoya, M., and N. Kumarakulasingam, 2017: The Lived Experience of Food Sovereignty: Gender, Indigenous Crops and Small-Scale Farming in Mtubatuba, South Africa. *J. Agrar. Chang.*, **17**, 480–496, doi:10.1111/joac.12170. <http://doi.wiley.com/10.1111/joac.12170> (Accessed May 10, 2018).
- Ngigi, M. W., U. Mueller, and R. Birner, 2017: Gender Differences in Climate Change Adaptation Strategies and Participation in Group-based Approaches: An Intra-household Analysis From Rural Kenya. *Ecol. Econ.*, doi:10.1016/j.ecolecon.2017.03.019.
- Nguyen, D., 2010: Evidence of the Impacts of Urban Sprawl on Social Capital. *Environ. Plan. B Plan. Des.*, **37**, 610–627, doi:10.1068/b35120. <http://journals.sagepub.com/doi/10.1068/b35120> (Accessed April 16, 2019).
- Niehof, A., 2004: The significance of diversification for rural livelihood systems. *Food Policy*, **29**, 321–338, doi:10.1016/j.foodpol.2004.07.009. <http://linkinghub.elsevier.com/retrieve/pii/S030691920400048X> (Accessed November 11, 2018).
- van Niekerk, J., and R. Wynberg, 2017a: Traditional seed and exchange systems cement social relations and provide a safety net: A case study from KwaZulu-Natal, South Africa. *Agroecol. Sustain. Food Syst.*, 1–25, doi:10.1080/21683565.2017.1359738. <https://www.tandfonline.com/doi/full/10.1080/21683565.2017.1359738> (Accessed June 4, 2018).
- , and ———, 2017b: Traditional seed and exchange systems cement social relations and provide a safety net: A case study from KwaZulu-Natal, South Africa. *Agroecol. Sustain. Food Syst.*, 1–25, doi:10.1080/21683565.2017.1359738. <https://www.tandfonline.com/doi/full/10.1080/21683565.2017.1359738> (Accessed April 15, 2019).
- Nightingale, A. J., 2017: Power and politics in climate change adaptation efforts: Struggles over authority and recognition in the context of political instability. *Geoforum*, **84**, 11–20,

- doi:10.1016/j.geoforum.2017.05.011.
<https://linkinghub.elsevier.com/retrieve/pii/S001671851730129X> (Accessed November 10, 2018).
- Nigussie, Z., and Coauthors, 2017: Factors influencing small-scale farmers' adoption of sustainable land management technologies in north-western Ethiopia. *Land use policy*, **67**, 57–64. <https://www.sciencedirect.com/science/article/pii/S0264837716305099> (Accessed November 9, 2018).
- del Ninno, C., P. A. Dorosh, and K. Subbarao, 2007: Food aid, domestic policy and food security: Contrasting experiences from South Asia and sub-Saharan Africa. *Food Policy*, **32**, 413–435, doi:10.1016/j.foodpol.2006.11.007. <http://linkinghub.elsevier.com/retrieve/pii/S0306919206001138> (Accessed November 11, 2018).
- Nizeyimana, E. L., G. W. Petersen, M. L. Imhoff, H. R. Sinclair, S. W. Waltman, D. S. Reed-Margetan, E. R. Levine, and J. M. Russo, 2001: Assessing the Impact of Land Conversion to Urban Use on Soils with Different Productivity Levels in the USA. *Soil Sci. Soc. Am. J.*, **65**, 391, doi:10.2136/sssaj2001.652391x. <https://www.soils.org/publications/sssaj/abstracts/65/2/391> (Accessed May 31, 2018).
- Noble, I., S. Huq, Y. Anokhin, J. Carmin, D. Goudou, F. Lansigan, B. Osman-Elasha, and A. Villamizar, 2014: Adaptation needs and options. *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, C. Field et al., Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 833–868.
- Noble, I. R., and Coauthors, 2015: Adaptation needs and options. *Clim. Chang. 2014 Impacts, Adapt. Vulnerability Part A Glob. Sect. Asp.*, 833–868, doi:10.1017/CBO9781107415379.019.
- North, M. P., S. L. Stephens, B. M. Collins, J. K. Agee, G. Aplet, J. F. Franklin, and P. Z. Fule, 2015a: Reform forest fire management. *Science (80-.)*, **349**, 1280–1281, doi:10.1126/science.aab2356. <http://www.sciencemag.org/cgi/doi/10.1126/science.aab2356> (Accessed April 18, 2019).
- North, M. P., S. L. Stephens, B. M. Collins, J. K. Agee, G. Aplet, J. F. Franklin, and P. Z. Fulé, 2015b: Reform forest fire management. *Science*, **349**, 1280–1281, doi:10.1126/science.aab2356. <http://www.ncbi.nlm.nih.gov/pubmed/26383934> (Accessed May 31, 2018).
- Norton, B. A., A. M. Coutts, S. J. Livesley, R. J. Harris, A. M. Hunter, and N. S. G. Williams, 2015: Planning for cooler cities: A framework to prioritise green infrastructure to mitigate high temperatures in urban landscapes. *Landsc. Urban Plan.*, **134**, 127–138, doi:10.1016/J.LANDURBPLAN.2014.10.018. <https://www.sciencedirect.com/science/article/pii/S0169204614002503> (Accessed June 2, 2018).
- Nowak, D. J., S. Hirabayashi, A. Bodine, and E. Greenfield, 2014: Tree and forest effects on air quality and human health in the United States. *Environ. Pollut.*, **193**, 119–129, doi:<https://doi.org/10.1016/j.envpol.2014.05.028>.
- Nunes, A., and Coauthors, 2016: Ecological restoration across the Mediterranean Basin as viewed by practitioners. *Sci. Total Environ.*, **566**, 722–732. <https://www.sciencedirect.com/science/article/pii/S004896971631066X> (Accessed November 11, 2018).
- Núñez, M., B. Civit, P. Muñoz, A. P. Arena, J. Rieradevall, and A. Antón, 2010: Assessing potential desertification environmental impact in life cycle assessment. *Int. J. Life Cycle Assess.*, **15**, 67–78, doi:10.1007/s11367-009-0126-0. <http://link.springer.com/10.1007/s11367-009-0126-0> (Accessed April 5, 2019).
- Nussbaum, M., and A. Sen, 1993: *The quality of life*.

- [https://books.google.com/books?hl=en&lr=&id=mOHnCwAAQBAJ&oi=fnd&pg=PR7&dq=Nussbaum,+M.,+%26+Sen,+A.+K.,+eds++\(1993\).+The+Quality+of+Life.+Oxford:+Oxford+University+Press.&ots=HJJRYVhaSz&sig=aBv5tCTrxpfHvzZwwMtrjBNI-Ns](https://books.google.com/books?hl=en&lr=&id=mOHnCwAAQBAJ&oi=fnd&pg=PR7&dq=Nussbaum,+M.,+%26+Sen,+A.+K.,+eds++(1993).+The+Quality+of+Life.+Oxford:+Oxford+University+Press.&ots=HJJRYVhaSz&sig=aBv5tCTrxpfHvzZwwMtrjBNI-Ns) (Accessed November 10, 2018).
- O'Mara, F., 2012a: The role of grasslands in food security and climate change. *Ann. Bot.*, **110**, 1263–1270. <https://academic.oup.com/aob/article-abstract/110/6/1263/112127> (Accessed November 12, 2018).
- O'Mara, F. P., 2012b: The role of grasslands in food security and climate change. *Ann. Bot.*, **110**, 1263–1270, doi:10.1093/aob/mcs209. <https://academic.oup.com/aob/article-lookup/doi/10.1093/aob/mcs209> (Accessed April 16, 2019).
- O'Neill, B. C., E. Kriegler, K. Riahi, K. L. Ebi, S. Hallegatte, T. R. Carter, R. Mathur, and D. P. van Vuuren, 2014a: A new scenario framework for climate change research: the concept of shared socioeconomic pathways. *Clim. Change*, **122**, 387–400, doi:10.1007/s10584-013-0905-2.
- O'Neill, B. C., E. Kriegler, K. Riahi, K. L. Ebi, S. Hallegatte, T. R. Carter, R. Mathur, and D. P. van Vuuren, 2014b: A new scenario framework for climate change research: the concept of shared socioeconomic pathways. *Clim. Change*, **122**, 387–400, doi:10.1007/s10584-013-0905-2. <https://doi.org/10.1007/s10584-013-0905-2>.
- O'Neill, B. C., E. Kriegler, K. Riahi, K. L. Ebi, S. Hallegatte, T. R. Carter, R. Mathur, and D. P. van Vuuren, 2014c: A new scenario framework for climate change research: the concept of shared socioeconomic pathways. *Clim. Change*, **122**, 387–400, doi:10.1007/s10584-013-0905-2. <http://link.springer.com/10.1007/s10584-013-0905-2> (Accessed November 10, 2018).
- , and Coauthors, 2017: The roads ahead: Narratives for shared socioeconomic pathways describing world futures in the 21st century. *Glob. Environ. Chang.*, **42**, 169–180, doi:10.1016/J.GLOENVCHA.2015.01.004. <https://www.sciencedirect.com/science/article/abs/pii/S0959378015000060> (Accessed April 19, 2019).
- Obersteiner, M., and Coauthors, 2016a: Assessing the land resource–food price nexus of the Sustainable Development Goals. *Sci. Adv.*, **2**, e1501499, doi:10.1126/sciadv.1501499. <http://advances.sciencemag.org/lookup/doi/10.1126/sciadv.1501499> (Accessed November 11, 2018).
- Obersteiner, M., and Coauthors, 2016b: Assessing the land resource–food price nexus of the Sustainable Development Goals. *Sci. Adv.*, doi:10.1126/sciadv.1501499.
- Odgaard, M. V., M. T. Knudsen, J. E. Hermansen, and T. Dalgaard, 2019: Targeted grassland production—A Danish case study on multiple benefits from converting cereal to grasslands for green biorefinery. *J. Clean. Prod.*,.
- Oehl, F., E. Laczko, H.-R. Oberholzer, J. Jansa, and S. Egli, 2017: Diversity and biogeography of arbuscular mycorrhizal fungi in agricultural soils. *Biol. Fertil. Soils*, **53**, 777–797, doi:10.1007/s00374-017-1217-x. <http://link.springer.com/10.1007/s00374-017-1217-x> (Accessed April 16, 2019).
- Oldeman, L., R. Hakkeling, and W. Sombroek, 1991: World map of the status of human-induced soil degradation: an explanatory note. Global Assessment of Soil Degradation (GLASOD). <http://agris.fao.org/agris-search/search.do?recordID=XF2015013172> (Accessed November 12, 2018).
- Olesen, J., and M. Bindi, 2002: Consequences of climate change for European agricultural productivity, land use and policy. *Eur. J. Agron.*, **16**, 239–262. <https://www.sciencedirect.com/science/article/pii/S1161030102000047> (Accessed May 30, 2018).

- de Oliveira Silva, R., L. G. Barioni, J. A. J. Hall, A. C. Moretti, R. Fonseca Veloso, P. Alexander, M. Crespolini, and D. Moran, 2017: Sustainable intensification of Brazilian livestock production through optimized pasture restoration. *Agric. Syst.*, **153**, 201–211, doi:10.1016/J.AGSY.2017.02.001. <https://www.sciencedirect.com/science/article/pii/S0308521X16303845> (Accessed June 2, 2018).
- , —, G. Queiroz Pellegrino, and D. Moran, 2018: The role of agricultural intensification in Brazil's Nationally Determined Contribution on emissions mitigation. *Agric. Syst.*, **161**, 102–112, doi:10.1016/J.AGSY.2018.01.003. <https://www.sciencedirect.com/science/article/pii/S0308521X17307655> (Accessed June 2, 2018).
- Oliver, C. D., N. T. Nassar, B. R. Lippke, and J. B. McCarter, 2014: Carbon, Fossil Fuel, and Biodiversity Mitigation With Wood and Forests. *J. Sustain. For.*, **33**, 248–275, doi:10.1080/10549811.2013.839386.
- Oliver, T. H., and M. D. Morecroft, 2014: Interactions between climate change and land use change on biodiversity: Attribution problems, risks, and opportunities. *Wiley Interdiscip. Rev. Clim. Chang.*, doi:10.1002/wcc.271.
- Oloo, J. O., and P. Omondi, 2017: Strengthening local institutions as avenues for climate change resilience. *Int. J. Disaster Resil. Built Environ.*, **8**, 573–588, doi:10.1108/IJDRBE-12-2013-0047. <http://www.emeraldinsight.com/doi/10.1108/IJDRBE-12-2013-0047> (Accessed November 10, 2018).
- Van Oost, K., and Coauthors, 2007: The impact of agricultural soil erosion on the global carbon cycle. *Science*, **318**, 626–629, doi:10.1126/science.1145724. <http://www.ncbi.nlm.nih.gov/pubmed/17962559> (Accessed April 12, 2019).
- Osbahr, H., C. Twyman, W. Neil Adger, and D. S. G. Thomas, 2008: Effective livelihood adaptation to climate change disturbance: Scale dimensions of practice in Mozambique. *Geoforum*, **39**, 1951–1964, doi:10.1016/j.geoforum.2008.07.010. <https://www.sciencedirect.com/science/article/pii/S0016718508001000> (Accessed November 11, 2018).
- Ostrom, E., 1990: An Institutional Approach to the Study of Self-organization and Self-governance in CPR Situations. *Gov. Commons Evol. institutions Collect. action*, doi:10.1017/CBO9781316423936.003.
- , 2000: Collective Action and the Evolution of Social Norms. *J. Econ. Perspect.*, **14**, 137–158, doi:10.1257/jep.14.3.137. <http://pubs.aeaweb.org/doi/10.1257/jep.14.3.137> (Accessed November 10, 2018).
- Osuri, A. M., and Coauthors, 2016: Contrasting effects of defaunation on aboveground carbon storage across the global tropics. *Nat. Commun.*, **7**, 11351, doi:10.1038/ncomms11351. <http://www.nature.com/articles/ncomms11351> (Accessed April 19, 2019).
- Overmars, K., E. Stehfest, A. Tabeau, H. van M.-L. U. Policy, and undefined 2014, Estimating the opportunity costs of reducing carbon dioxide emissions via avoided deforestation, using integrated assessment modelling. *Elsevier.*, <https://www.sciencedirect.com/science/article/pii/S0264837714000799> (Accessed April 18, 2019).
- Overmars, K. P., E. Stehfest, A. Tabeau, H. van Meijl, A. M. Beltrán, and T. Kram, 2014: Estimating the opportunity costs of reducing carbon dioxide emissions via avoided deforestation, using integrated assessment modelling. *Land use policy*, **41**, 45–60, doi:10.1016/J.LANDUSEPOL.2014.04.015. <https://www.sciencedirect.com/science/article/pii/S0264837714000799> (Accessed June 1, 2018).
- Pacala, S., and R. Socolow, 2004: Stabilization Wedges: Solving the Climate Problem for the Next 50

- Years with Current Technologies. *Science* (80-), **305**, 968 LP-972. <http://science.sciencemag.org/content/305/5686/968.abstract>.
- Padgham, J., J. Jabbour, and K. Dietrich, 2014: Managing change and building resilience: A multi-stressor analysis of urban and peri-urban agriculture in Africa and Asia. *Urban Clim.*, **1**, 183–204. <https://www.sciencedirect.com/science/article/pii/S2212095515000139> (Accessed May 1, 2018).
- Le Page, Y., and Coauthors, 2013a: Sensitivity of climate mitigation strategies to natural disturbances. *Environ. Res. Lett.*, **8**, 15018, doi:10.1088/1748-9326/8/1/015018. <http://stacks.iop.org/1748-9326/8/i=1/a=015018?key=crossref.0f18e6b8532829147cdb06fa69f8d38d> (Accessed November 11, 2018).
- Le Page, Y., and Coauthors, 2013b: Sensitivity of climate mitigation strategies to natural disturbances. *Environ. Res. Lett.*, **8**, 15018, doi:10.1088/1748-9326/8/1/015018. <http://stacks.iop.org/1748-9326/8/i=1/a=015018?key=crossref.0f18e6b8532829147cdb06fa69f8d38d> (Accessed November 11, 2018).
- Pahl-Wostl, C., A. Bhaduri, and A. Bruns, 2018: Editorial special issue: The Nexus of water, energy and food – An environmental governance perspective. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2018.06.021.
- Palacios, M., E. Huber-Sannwald, L. B.-L. U. Policy, and undefined 2013, Landscape diversity in a rural territory: emerging land use mosaics coupled to livelihood diversification. *Elsevier*.
- , ——, L. Barrios, F. de Paz, J. Hernández, and M. Mendoza, 2013: Landscape diversity in a rural territory: emerging land use mosaics coupled to livelihood diversification. *Land use policy*, **30**, 814–824. <https://www.sciencedirect.com/science/article/pii/S0264837712001123> (Accessed November 11, 2018).
- Palm, C., H. Blanco-Canqui, F. DeClerck, L. Gatere, and P. Grace, 2014: Conservation agriculture and ecosystem services: An overview. *Agric. Ecosyst. Environ.*, **187**, 87–105, doi:10.1016/J.AGEE.2013.10.010. <https://www.sciencedirect.com/science/article/pii/S0167880913003502> (Accessed June 5, 2018).
- Pan, G., P. Smith, and W. Pan, 2009: The role of soil organic matter in maintaining the productivity and yield stability of cereals in China. *Agric. Ecosyst. Environ.*, **129**, 344–348, doi:10.1016/J.AGEE.2008.10.008. <https://www.sciencedirect.com/science/article/pii/S0167880908002703> (Accessed November 9, 2018).
- Papargyropoulou, E., R. Lozano, J. K. Steinberger, N. Wright, and Z. bin Ujang, 2014: The food waste hierarchy as a framework for the management of food surplus and food waste. *J. Clean. Prod.*, **76**, 106–115, doi:10.1016/J.JCLEPRO.2014.04.020. <https://www.sciencedirect.com/science/article/pii/S0959652614003680> (Accessed April 16, 2019).
- Parfitt, J., M. Barthel, and S. Macnaughton, 2010: Food waste within food supply chains: quantification and potential for change to 2050. *Philos. Trans. R. Soc. B Biol. Sci.*, **365**, 3065–3081, doi:10.1098/rstb.2010.0126. <http://www.royalsocietypublishing.org/doi/10.1098/rstb.2010.0126> (Accessed April 16, 2019).
- Parizeau, K., M. von Massow, and R. Martin, 2015: Household-level dynamics of food waste production and related beliefs, attitudes, and behaviours in Guelph, Ontario. *Waste Manag.*, **35**, 207–217, doi:10.1016/J.WASMAN.2014.09.019. <https://www.sciencedirect.com/science/article/pii/S0956053X14004413> (Accessed April 16, 2019).
- Park, C. M. Y., B. White, and Julia, 2015: We are not all the same: taking gender seriously in food sovereignty discourse. *Third World Q.*, **36**, 584–599, doi:10.1080/01436597.2015.1002988.

- <http://www.tandfonline.com/doi/full/10.1080/01436597.2015.1002988> (Accessed April 16, 2019).
- Parkinson, S., and Coauthors, 2019: Balancing clean water-climate change mitigation trade-offs. *Environ. Res. Lett.*, **14**, 14009, doi:10.1088/1748-9326/aaf2a3. <http://stacks.iop.org/1748-9326/14/i=1/a=014009?key=crossref.67da0e6d5ac6e628a3c2f279956bc029> (Accessed April 14, 2019).
- Parnell, S., D. Simon, and C. Vogel, 2007: Global environmental change: conceptualising the growing challenge for cities in poor countries. *Area*, **39**, 357–369, doi:10.1111/j.1475-4762.2007.00760.x. <http://doi.wiley.com/10.1111/j.1475-4762.2007.00760.x> (Accessed April 16, 2019).
- Parr, T. W., A. R. J. Sier, R. W. Battarbee, A. Mackay, and J. Burgess, 2003: Detecting environmental change: science and society—perspectives on long-term research and monitoring in the 21st century. *Sci. Total Environ.*, **310**, 1–8, doi:10.1016/S0048-9697(03)00257-2. <https://www.sciencedirect.com/science/article/pii/S0048969703002572> (Accessed April 16, 2019).
- Pascual, U., and Coauthors, 2017: Valuing nature’s contributions to people: the IPBES approach. *Curr. Opin. Environ. Sustain.*, **26–27**, 7–16, doi:10.1016/J.COSUST.2016.12.006. <https://www.sciencedirect.com/science/article/pii/S1877343517300040> (Accessed April 5, 2019).
- Paterson, R. R. M., L. Kumar, F. Shabani, and N. Lima, 2017: World climate suitability projections to 2050 and 2100 for growing oil palm. *J. Agric. Sci.*, **155**, 689–702, doi:10.1017/S0021859616000605. https://www.cambridge.org/core/product/identifier/S0021859616000605/type/journal_article (Accessed April 4, 2019).
- Patnaik, A., J. Jongerden, and G. Ruivenkamp, 2017: Repossession through sharing of and access to seeds: different cases and practices. *Int. Rev. Sociol.*, **27**, 179–201, doi:10.1080/03906701.2016.1235213. <https://www.tandfonline.com/doi/full/10.1080/03906701.2016.1235213> (Accessed June 1, 2018).
- Patrizio, P., and Coauthors, 2018: Reducing US Coal Emissions Can Boost Employment. *Joule*, **2**, 2633–2648.
- Patt, A., P. Suarez, and U. Hess, 2010: How do small-holder farmers understand insurance, and how much do they want it? Evidence from Africa. *Glob. Environ. Chang.*, **20**, 153–161. <https://www.sciencedirect.com/science/article/pii/S0959378009000922> (Accessed November 11, 2018).
- Paustian, K., J. Lehmann, S. Ogle, D. Reay, G. P. Robertson, and P. Smith, 2016: Climate-smart soils. *Nature*, **532**, 49–57, doi:10.1038/nature17174. <http://www.nature.com/doi/10.1038/nature17174> (Accessed June 2, 2018).
- Pawson, S. M., A. Brin, E. G. Brockerhoff, D. Lamb, T. W. Payn, A. Paquette, and J. A. Parrotta, 2013: Plantation forests, climate change and biodiversity. *Biodivers. Conserv.*, **22**, 1203–1227, doi:10.1007/s10531-013-0458-8. <http://link.springer.com/10.1007/s10531-013-0458-8> (Accessed April 4, 2019).
- Payn, T., and Coauthors, 2015: Changes in planted forests and future global implications. *For. Ecol. Manage.*, **352**, 57–67. <https://www.sciencedirect.com/science/article/pii/S0378112715003473> (Accessed November 11, 2018).
- Pedercini, M., G. Zuellich, K. Dianati, and S. Arquitt, 2018: Toward achieving Sustainable Development Goals in Ivory Coast: Simulating pathways to sustainable development. *Sustain. Dev.*, **0**, doi:10.1002/sd.1721. <https://doi.org/10.1002/sd.1721>.
- Pellegrini, L., and L. Tasciotti, 2014: Crop diversification, dietary diversity and agricultural income: Empirical evidence from eight developing countries. *Can. J. Dev. Stud.*, **35**, 211–227,

- doi:10.1080/02255189.2014.898580.
<http://www.tandfonline.com/doi/abs/10.1080/02255189.2014.898580> (Accessed May 31, 2018).
- Pelletier, J., N. Gélinas, M. Skutsch, J. Pelletier, N. Gélinas, and M. Skutsch, 2016: The Place of Community Forest Management in the REDD+ Landscape. *Forests*, **7**, 170, doi:10.3390/f7080170. <http://www.mdpi.com/1999-4907/7/8/170> (Accessed April 5, 2019).
- Pendleton, L. *et al.* Estimating Global 'Blue Carbon' Emissions from Conversion and Degradation of Vegetated Coastal Ecosystems. *PLoS One* **7**, (2012).
- Pereira, H., P. Leadley, ... V. P., and undefined 2010, Scenarios for global biodiversity in the 21st century. *science.sciencemag.org*,. <http://science.sciencemag.org/content/330/6010/1496.short> (Accessed November 11, 2018).
- Pereira, H. M., and Coauthors, 2010: Scenarios for Global Biodiversity in the 21st Century. *Science* (80-.), **330**, 1496–1502. <http://api.iucnredlist.org/go/panthera-leo>.
- Perugini, L., L. Caporaso, S. Marconi, A. Cescatti, B. Quesada, N. De Noblet-Ducoudré, J. I. House, and A. Arneth, 2017: Biophysical effects on temperature and precipitation due to land cover change. *Environ. Res. Lett.*, doi:10.1088/1748-9326/aa6b3f.
- Petersen, A. K., and B. Solberg, 2005: Environmental and economic impacts of substitution between wood products and alternative materials: a review of micro-level analyses from Norway and Sweden. *For. Policy Econ.*, **7**, 249–259, doi:10.1016/S1389-9341(03)00063-7. <https://www.sciencedirect.com/science/article/pii/S1389934103000637> (Accessed April 16, 2019).
- Peterson, G. D., and Coauthors, 2018: Welcoming different perspectives in IPBES: “Nature’s contributions to people” and “Ecosystem services” *Ecol. Soc.*, **23**, art39, doi:10.5751/ES-10134-230139. <https://www.ecologyandsociety.org/vol23/iss1/art39/> (Accessed April 5, 2019).
- Peterson, N. D., 2012: Developing Climate Adaptation: The Intersection of Climate Research and Development Programmes in Index Insurance. *Dev. Change*, doi:10.1111/j.1467-7660.2012.01767.x.
- Phelps, J., E. Webb, and A. Agrawal, 2010: Does REDD+ threaten to recentralize forest governance? *Science* (80-.), **328**, 312–313. <http://science.sciencemag.org/content/328/5976/312.short> (Accessed November 10, 2018).
- Pikaar, I., and Coauthors, 2018: Decoupling Livestock from Land Use through Industrial Feed Production Pathways. *Environ. Sci. Technol.*, **52**, 7351–7359, doi:10.1021/acs.est.8b00216. <http://pubs.acs.org/doi/10.1021/acs.est.8b00216> (Accessed April 14, 2019).
- Pimentel, D., 2006: Soil Erosion: A Food and Environmental Threat. *Environ. Dev. Sustain.*, **8**, 119–137, doi:10.1007/s10668-005-1262-8. <http://link.springer.com/10.1007/s10668-005-1262-8> (Accessed November 11, 2018).
- Pimentel, D., R. Zuniga, and D. Morrison, 2005: Update on the environmental and economic costs associated with alien-invasive species in the United States. *Ecol. Econ.*, **52**, 273–288. <https://www.sciencedirect.com/science/article/pii/S0921800904003027> (Accessed November 11, 2018).
- Platteau, J.-P., O. De Bock, and W. Gelade, 2017: The Demand for Microinsurance: A Literature Review. *World Dev.*, **94**, 139–156, doi:10.1016/j.worlddev.2017.01.010. <https://linkinghub.elsevier.com/retrieve/pii/S0305750X1730013X> (Accessed November 11, 2018).
- Poepplau, C., and A. Don, 2015: Carbon sequestration in agricultural soils via cultivation of cover crops – A meta-analysis. *Agric. Ecosyst. Environ.*, **200**, 33–41, doi:10.1016/J.AGEE.2014.10.024.

- <https://www.sciencedirect.com/science/article/pii/S0167880914004873> (Accessed May 31, 2018).
- POEPLAU, C., A. DON, L. VESTERDAL, J. LEIFELD, B. VAN WESEMAEL, J. SCHUMACHER, and A. GENSIOR, 2011: Temporal dynamics of soil organic carbon after land-use change in the temperate zone - carbon response functions as a model approach. *Glob. Chang. Biol.*, **17**, 2415–2427, doi:10.1111/j.1365-2486.2011.02408.x. <http://doi.wiley.com/10.1111/j.1365-2486.2011.02408.x> (Accessed November 11, 2018).
- Pohnan, E., H. Ompusunggu, and C. Webb, 2015: Does Tree Planting Change Minds? Assessing the Use of Community Participation in Reforestation to Address Illegal Logging in West Kalimantan. *Trop. Conserv. Sci.*, **8**, 45–57, doi:10.1177/194008291500800107. <http://journals.sagepub.com/doi/10.1177/194008291500800107> (Accessed June 2, 2018).
- Poore, J., T. Nemecek 2018. Reducing food's environmental impacts through producers and consumers. *Science* **360**, 987–992.
- Popkin, B. M., 2008: Will China's Nutrition Transition Overwhelm Its Health Care System And Slow Economic Growth? *Health Aff.*, **27**, 1064–1076, doi:10.1377/hlthaff.27.4.1064. <http://www.healthaffairs.org/doi/10.1377/hlthaff.27.4.1064> (Accessed April 16, 2019).
- Popp, A., H. Lotze-campen, and B. Bodirsky, 2010: Food consumption , diet shifts and associated non-CO 2 greenhouse gases from agricultural production. *Glob. Environ. Chang.*, **20**, 451–462, doi:10.1016/j.gloenvcha.2010.02.001. <http://dx.doi.org/10.1016/j.gloenvcha.2010.02.001>.
- Popp, A., and Coauthors, 2011a: The economic potential of bioenergy for climate change mitigation with special attention given to implications for the land system. *Environ. Res. Lett.*, **6**, doi:10.1088/1748-9326/6/3/034017.
- , and Coauthors, 2011b: The economic potential of bioenergy for climate change mitigation with special attention given to implications for the land system. *Environ. Res. Lett.*, **6**, 34017, doi:10.1088/1748-9326/6/3/034017. <http://stacks.iop.org/1748-9326/6/i=3/a=034017?key=crossref.a2656c08649c5f8b7418b37f3dd063a4> (Accessed November 11, 2018).
- , H. Lotze-Campen, M. Leimbach, B. Knopf, T. Beringer, N. Bauer, and B. Bodirsky, 2011c: On sustainability of bioenergy production: Integrating co-emissions from agricultural intensification. *Biomass and Bioenergy*, **35**, 4770–4780, doi:https://doi.org/10.1016/j.biombioe.2010.06.014. <http://www.sciencedirect.com/science/article/pii/S0961953410002230>.
- , and Coauthors, 2014a: Land-use transition for bioenergy and climate stabilization: model comparison of drivers, impacts and interactions with other land use based mitigation options. *Clim. Change*, **123**, 495–509, doi:10.1007/s10584-013-0926-x. <http://link.springer.com/10.1007/s10584-013-0926-x> (Accessed November 11, 2018).
- , and Coauthors, 2014b: Land-use transition for bioenergy and climate stabilization: Model comparison of drivers, impacts and interactions with other land use based mitigation options. *Clim. Change*, doi:10.1007/s10584-013-0926-x.
- , and Coauthors, 2017: Land-use futures in the shared socio-economic pathways. *Glob. Environ. Chang.*, **42**, 331–345, doi:10.1016/j.gloenvcha.2016.10.002. <http://dx.doi.org/10.1016/j.gloenvcha.2016.10.002>.
- Porter, J. R., L. Xie, A. J. Challinor, K. Cochrane, S. M. Howden, M. M. Iqbal, D. B. Lobell, and M. I. Travasso, 2014: Food security and food production systems. *Climate Change 2014: Impacts*,

- Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, Cambridge University Press
https://scholar.google.com/scholar?hl=en&as_sdt=0%2C5&q=Porter%2C+J.+R.%2C+L.+Xie%2C+A.+J.+Challinor%2C+K.+Cochrane%2C+S.+M.+Howden%2C+M.+M.+Iqbal%2C+D.+B.+Lobell%2C+and+M.+I.+Travasso%2C+2014%3A+Chapter+7%3A+Food+security+and+food+production+systems.+Cam (Accessed November 7, 2018).
- Porter, S. D., D. S. Reay, P. Higgins, and E. Bomberg, 2016a: A half-century of production-phase greenhouse gas emissions from food loss & waste in the global food supply chain. *Sci. Total Environ.*, **571**, 721–729, doi:10.1016/j.scitotenv.2016.07.041.
- , ———, ———, and ———, 2016b: A half-century of production-phase greenhouse gas emissions from food loss & waste in the global food supply chain. *Sci. Total Environ.*, **571**, 721–729, doi:10.1016/j.scitotenv.2016.07.041.
<https://www.sciencedirect.com/science/article/pii/S0048969716314863> (Accessed June 7, 2018).
- Poteete, A. R., and E. Ostrom, 2004: In pursuit of comparable concepts and data about collective action. *Agric. Syst.*, doi:10.1016/j.agsy.2004.07.002.
- Pothukuchi, K., and J. L. Kaufman, 1999: Placing the food system on the urban agenda: The role of municipal institutions in food systems planning. *Agric. Human Values*, **16**, 213–224, doi:10.1023/A:1007558805953. <http://link.springer.com/10.1023/A:1007558805953> (Accessed April 16, 2019).
- Potschin, M. B., and R. H. Haines-Young, 2011: Ecosystem services. *Prog. Phys. Geogr. Earth Environ.*, **35**, 575–594, doi:10.1177/0309133311423172. <http://journals.sagepub.com/doi/10.1177/0309133311423172> (Accessed April 5, 2019).
- Potts, S. G., J. C. Biesmeijer, C. Kremen, P. Neumann, O. Schweiger, and W. E. Kunin, 2010: Global pollinator declines: Trends, impacts and drivers. *Trends Ecol. Evol.*, doi:10.1016/j.tree.2010.01.007.
- Poulton, C., J. Kydd, S. Wiggins, and A. Dorward, 2006: State intervention for food price stabilisation in Africa: Can it work? *Food Policy*, **31**, 342–356, doi:10.1016/J.FOODPOL.2006.02.004. <https://www.sciencedirect.com/science/article/pii/S0306919206000261> (Accessed June 1, 2018).
- Powell, J., 1999: Race, poverty, and urban sprawl: Access to opportunities through regional strategies. *Forum Soc. Econ.*, **28**, 1–20, doi:10.1007/BF02833980. <http://www.tandfonline.com/doi/abs/10.1007/BF02833980> (Accessed April 16, 2019).
- Powell, T. W. R. & Lenton, T. M. Future carbon dioxide removal via biomass energy constrained by agricultural efficiency and dietary trends. *Energy and Environmental Science* (2012).
- Powers, R. P., and W. Jetz, 2019: Global habitat loss and extinction risk of terrestrial vertebrates under future land-use-change scenarios. *Nat. Clim. Chang.*, **9**, 323–329, doi:10.1038/s41558-019-0406-z. <http://www.nature.com/articles/s41558-019-0406-z> (Accessed April 13, 2019).
- Powlson, D. S., C. M. Stirling, M. L. Jat, B. G. Gerard, C. A. Palm, P. A. Sanchez, and K. G. Cassman, 2014: Limited potential of no-till agriculture for climate change mitigation. *Nat. Clim. Chang.*, **4**, 678–683, doi:10.1038/nclimate2292. <http://www.nature.com/doi/10.1038/nclimate2292> (Accessed June 5, 2018).
- , C. M. Stirling, C. Thierfelder, R. P. White, and M. L. Jat, 2016: Does conservation agriculture deliver climate change mitigation through soil carbon sequestration in tropical agro-ecosystems? *Agric. Ecosyst. Environ.*, **220**, 164–174, doi:10.1016/j.agee.2016.01.005. <https://www.sciencedirect.com/science/article/pii/S0167880916300056> (Accessed June 5, 2018).

- Pozzi, W., and Coauthors, 2013: Toward Global Drought Early Warning Capability: Expanding International Cooperation for the Development of a Framework for Monitoring and Forecasting. *Bull. Am. Meteorol. Soc.*, **94**, 776–785, doi:10.1175/BAMS-D-11-00176.1. <http://journals.ametsoc.org/doi/abs/10.1175/BAMS-D-11-00176.1> (Accessed November 11, 2018).
- Pradhan P, Reusser DE, Kropp JP (2013) Embodied Greenhouse Gas Emissions in Diets. *PLOS ONE* 8(5): e62228. <https://doi.org/10.1371/journal.pone.0062228>
- Prathapar, S. A., 1988: *How to Manage Salinity in Irrigated Lands: A Selective Review with Particular Reference to Irrigation in Developing Countries*.
- Pratt, K. & Moran, D. Evaluating the cost-effectiveness of global biochar mitigation potential. *Biomass and Bioenergy* (2010).
- Pretty, J., 2003: Social Capital and the Collective Management of Resources. *Science* (80-.), **302**, 1912–1914, doi:10.1126/science.1090847. <http://www.sciencemag.org/cgi/doi/10.1126/science.1090847> (Accessed November 10, 2018).
- Pretty, J., and Z. P. Z. Bharucha, 2014: Sustainable intensification in agricultural systems. *Ann. Bot.*, **114**, 1571–1596, doi:10.1093/aob/mcu205. <https://academic.oup.com/aob/article-abstract/114/8/1571/210078> (Accessed May 30, 2018).
- Pretty, J., and Coauthors, 2018: Global assessment of agricultural system redesign for sustainable intensification. *Nat. Sustain.*, **1**, 441.
- Putz, F. E., and Coauthors, 2012: Sustaining conservation values in selectively logged tropical forests: The attained and the attainable. *Conserv. Lett.*, **5**, 296–303, doi:10.1111/j.1755-263X.2012.00242.x.
- Qadir, M., A. D. Noble, and C. Chartres, 2013: Adapting to climate change by improving water productivity of soils in dry areas. *L. Degrad. Dev.*, **24**, 12–21, doi:10.1002/ldr.1091. <http://doi.wiley.com/10.1002/ldr.1091> (Accessed November 11, 2018).
- Qian, J., Y. Peng, C. Luo, C. Wu, and Q. Du, 2015: Urban Land Expansion and Sustainable Land Use Policy in Shenzhen: A Case Study of China's Rapid Urbanization. *Sustainability*, **8**, 16, doi:10.3390/su8010016. <http://www.mdpi.com/2071-1050/8/1/16> (Accessed May 31, 2018).
- Qin, Z., J. B. Dunn, H. Kwon, S. Mueller, and M. M. Wander, 2016: Soil carbon sequestration and land use change associated with biofuel production: Empirical evidence. *GCB Bioenergy*, doi:10.1111/gcbb.12237.
- Rahman, M. R., and S. H. Bulbul, 2015: Adoption of Water Saving Irrigation Techniques for Sustainable Rice Production in Bangladesh. *Environ. Ecol. Res.*, **3**, 1–8, doi:10.13189/EER.2015.030101. http://www.hrpub.org/journals/article_info.php?aid=2169 (Accessed April 12, 2019).
- Rahman, S., 2010: Women's Labour Contribution to Productivity and Efficiency in Agriculture: Empirical Evidence From Bangladesh. *J. Agric. Econ.*, **61**, 318–342, doi:10.1111/j.1477-9552.2010.00243.x. <http://doi.wiley.com/10.1111/j.1477-9552.2010.00243.x> (Accessed April 16, 2019).
- Rakodi, C., 1999: A Capital Assets Framework for Analysing Household Livelihood Strategies: Implications for Policy. *Dev. Policy Rev.*, **17**, 315–342, doi:10.1111/1467-7679.00090. <http://doi.wiley.com/10.1111/1467-7679.00090> (Accessed November 11, 2018).
- Raleigh, C., H. J. Choi, and D. Kniveton, 2015a: The devil is in the details: An investigation of the relationships between conflict, food price and climate across Africa. *Glob. Environ. Chang.*, **32**, 187–199, doi:10.1016/j.gloenvcha.2015.03.005.

- <https://linkinghub.elsevier.com/retrieve/pii/S0959378015000357> (Accessed November 11, 2018).
- , ——, and ——, 2015b: The devil is in the details: An investigation of the relationships between conflict, food price and climate across Africa. *Glob. Environ. Chang.*, **32**, 187–199, doi:10.1016/j.gloenvcha.2015.03.005. <https://www.sciencedirect.com/science/article/pii/S0959378015000357> (Accessed November 11, 2018).
- Ramage, M. H., and Coauthors, 2017a: The wood from the trees: The use of timber in construction. *Renew. Sustain. Energy Rev.*, **68**, 333–359, doi:10.1016/J.RSER.2016.09.107. <https://www.sciencedirect.com/science/article/pii/S1364032116306050> (Accessed April 5, 2019).
- , and Coauthors, 2017b: The wood from the trees: The use of timber in construction. *Renew. Sustain. Energy Rev.*, **68**, 333–359, doi:10.1016/j.rser.2016.09.107. <https://linkinghub.elsevier.com/retrieve/pii/S1364032116306050> (Accessed April 18, 2019).
- Ramanathan, V., P. J. Crutzen, J. T. Kiehl, and D. Rosenfeld, 2001: Aerosols, Climate, and the Hydrological Cycle. *Science*, **294**, 2119–2124, doi:10.1126/science.250.4988.1669. http://science.sciencemag.org/content/294/5549/2119?casa_token=qGzy4pUeoCAAAAAA:uyXpj1AV3hc5IwMTBpX1H8cCVUIXkKd4ZKpgrLkkgOBO269UHWkik23tHUctkW8V29Biqx7lJtu0JYk (Accessed August 11, 2018).
- Rametsteiner, E., and M. Simula, 2003: Forest certification—an instrument to promote sustainable forest management? *J. Environ. Manage.*, **67**, 87–98, doi:10.1016/S0301-4797(02)00191-3. <https://www.sciencedirect.com/science/article/pii/S0301479702001913> (Accessed April 5, 2019).
- Randerson, J. T., Y. Chen, G. R. van der Werf, B. M. Rogers, and D. C. Morton, 2012: Global burned area and biomass burning emissions from small fires. *J. Geophys. Res. Biogeosciences*, **117**, n/a-n/a, doi:10.1029/2012JG002128. <http://doi.wiley.com/10.1029/2012JG002128> (Accessed November 12, 2018).
- Rao, S., and Coauthors, 2017: Future air pollution in the Shared Socio-economic Pathways. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2016.05.012.
- Rawlins, A., J. M.- Geoderma, and undefined 2010, Social and economic aspects of peatland management in Northern Europe, with particular reference to the English case. *Elsevier.*,
- Redford, K. H., and W. M. Adams, 2009: Payment for Ecosystem Services and the Challenge of Saving Nature. *Conserv. Biol.*, **23**, 785–787, doi:10.1111/j.1523-1739.2009.01271.x. <http://doi.wiley.com/10.1111/j.1523-1739.2009.01271.x> (Accessed April 5, 2019).
- Regmi, A., and B. Meade, 2013: Demand side drivers of global food security. *Glob. Food Sec.*, **2**, 166–171, doi:10.1016/j.gfs.2013.08.001. <https://linkinghub.elsevier.com/retrieve/pii/S2211912413000369> (Accessed November 11, 2018).
- Reichardt, M., C. Jürgens, U. Klöble, J. Hüter, and K. Moser, 2009a: Dissemination of precision farming in Germany: acceptance, adoption, obstacles, knowledge transfer and training activities. *Precis. Agric.*, **10**, 525–545, doi:10.1007/s11119-009-9112-6. <http://link.springer.com/10.1007/s11119-009-9112-6> (Accessed November 11, 2018).
- , ——, ——, ——, and ——, 2009b: Dissemination of precision farming in Germany: acceptance, adoption, obstacles, knowledge transfer and training activities. *Precis. Agric.*, **10**, 525–545, doi:10.1007/s11119-009-9112-6. <http://link.springer.com/10.1007/s11119-009-9112-6> (Accessed April 18, 2019).
- Reidsma, P., F. Ewert, A. O. Lansink, and R. Leemans, 2010: Adaptation to climate change and climate variability in European agriculture: The importance of farm level responses. *Eur. J. Agron.*, doi:10.1016/j.eja.2009.06.003.

- Reilly, J., and Coauthors, 2012a: Using land to mitigate climate change: Hitting the target, recognizing the trade-offs. *Environ. Sci. Technol.*, doi:10.1021/es2034729.
- , and Coauthors, 2012b: Using Land To Mitigate Climate Change: Hitting the Target, Recognizing the Trade-offs. *Environ. Sci. Technol.*, **46**, 5672–5679, doi:10.1021/es2034729. <http://pubs.acs.org/doi/10.1021/es2034729> (Accessed November 11, 2018).
- Reisman, E., 2017a: Troubling Tradition, Community, and Self-Reliance: Reframing Expectations for Village Seed Banks. *World Dev.*, **98**, 160–168, doi:10.1016/J.WORLDDEV.2017.04.024. <https://www.sciencedirect.com/science/article/pii/S0305750X17301353> (Accessed June 1, 2018).
- , 2017b: Troubling Tradition, Community, and Self-Reliance: Reframing Expectations for Village Seed Banks. *World Dev.*, **98**, 160–168, doi:10.1016/j.worlddev.2017.04.024. <https://linkinghub.elsevier.com/retrieve/pii/S0305750X17301353> (Accessed April 18, 2019).
- Renforth, P., W. M. Mayes, A. P. Jarvis, I. T. Burke, D. A. C. Manning, and K. Gruiz, 2012: Contaminant mobility and carbon sequestration downstream of the Ajka (Hungary) red mud spill: The effects of gypsum dosing. *Sci. Total Environ.*, **421–422**, 253–259, doi:10.1016/J.SCITOTENV.2012.01.046. <https://www.sciencedirect.com/science/article/pii/S0048969712000897> (Accessed June 7, 2018).
- Rengasamy, P., 2006: World salinization with emphasis on Australia. *J. Exp. Bot.*, **57**, 1017–1023, doi:10.1093/jxb/erj108.
- Revi, A., D. E. Satterthwaite, F. Aragón-Durand, J. Corfee-Morlot, R. B. R. Kiunsi, M. Pelling, Roberts, D.C., and W. Solecki, 2014: Urban Areas. *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*.
- Rey Benayas, J. M., A. C. Newton, A. Diaz, and J. M. Bullock, 2009: Enhancement of biodiversity and ecosystem services by ecological restoration: A meta-analysis. *Science (80-)*, doi:10.1126/science.1172460.
- Reyer, C., M. Guericke, and P. L. Ibsch, 2009: Climate change mitigation via afforestation, reforestation and deforestation avoidance: And what about adaptation to environmental change? *New For.*, **38**, 15–34, doi:10.1007/s11056-008-9129-0. <http://link.springer.com/10.1007/s11056-008-9129-0> (Accessed June 1, 2018).
- R. Hijbeek, M.K. van Ittersum, H.F.M. ten Berge, G. Gort, H. Spiegel, A.P. Whitmore 2017 Do organic inputs matter – a meta-analysis of additional yield effects for arable crops in Europe. *Plant Soil* (2017) 411: 293. <https://doi.org/10.1007/s11104-016-3031-x>
- Riahi, K., and Coauthors, 2017a: The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: An overview. *Glob. Environ. Chang.*, **42**, 153–168, doi:10.1016/j.gloenvcha.2016.05.009.
- , and Coauthors, 2017b: The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: An overview. *Glob. Environ. Chang.*, **42**, 153–168, doi:10.1016/J.GLOENVCHA.2016.05.009. <http://www.sciencedirect.com/science/article/pii/S0959378016300681> (Accessed December 24, 2017).
- , and Coauthors, 2017c: The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: An overview. *Glob. Environ. Chang.*, **42**, 153–168, doi:10.1016/j.gloenvcha.2016.05.009. <https://linkinghub.elsevier.com/retrieve/pii/S0959378016300681> (Accessed November 11, 2018).
- Ricci, L., 2012: Peri-Urban Livelihood and Adaptive Capacity: Urban Development in Dar Es Salaam. *Cons. J. Sustain. Dev.*, **7**, 46–63.

- Richards, M., and Coauthors, 2017: Highresolution spatial modelling of greenhouse gas emissions from landuse change to energy crops in the United Kingdom. *GCB Bioenergy*, doi:10.1111/gcbb.12360.
- Ridoutt, B., P. Sanguansri, L. Bonney, S. Crimp, G. L.- Climate, and undefined 2016, Climate change adaptation strategy in the food industry—insights from product carbon and water footprints. *mdpi.com.*, <https://www.mdpi.com/2225-1154/4/2/26html> (Accessed November 8, 2018).
- Rigg, J., 2006: Land, farming, livelihoods, and poverty: Rethinking the links in the Rural South. *World Dev.*, **34**, 180–202, doi:10.1016/j.worlddev.2005.07.015. <http://linkinghub.elsevier.com/retrieve/pii/S0305750X05001907> (Accessed November 11, 2018).
- Ringler, C., and R. Lawford, 2013: The nexus across water, energy, land and food (WELF): potential for improved resource use efficiency? *Curr. Opin. Environ. Sustain.*, **5**, 617–624, doi:10.1016/J.COSUST.2013.11.002. <https://www.sciencedirect.com/science/article/pii/S1877343513001504> (Accessed April 16, 2019).
- , D. Willenbockel, N. Perez, M. Rosegrant, T. Zhu, and N. Matthews, 2016: Global linkages among energy, food and water: an economic assessment. *J. Environ. Stud. Sci.*, **6**, 161–171, doi:10.1007/s13412-016-0386-5.
- Ritzema, R. S., and Coauthors, 2017: Is production intensification likely to make farm households food-adequate? A simple food availability analysis across smallholder farming systems from East and West Africa. *Food Secur.*, **9**, 115–131, doi:10.1007/s12571-016-0638-y. <http://link.springer.com/10.1007/s12571-016-0638-y> (Accessed May 10, 2018).
- Rivera-Ferre, M. G., F. López-i-Gelats, M. Howden, P. Smith, J. F. Morton, and M. Herrero, 2016: Reframing the climate change debate in the livestock sector: mitigation and adaptation options. *Wiley Interdiscip. Rev. Clim. Chang.*, **7**, 869–892, doi:10.1002/wcc.421. <http://doi.wiley.com/10.1002/wcc.421> (Accessed April 4, 2019).
- Robert, M., A. Thomas, M. Sekhar, S. Badiger, L. Ruiz, M. Willaume, D. Leenhardt, and J.-E. Bergez, 2017: Farm Typology in the Berambadi Watershed (India): Farming Systems Are Determined by Farm Size and Access to Groundwater. *Water*, **9**, 51, doi:10.3390/w9010051. <http://www.mdpi.com/2073-4441/9/1/51> (Accessed June 2, 2018).
- Roberts, K. G., Gloy, B. A., Joseph, S., Scott, N. R. & Lehmann, J. Life cycle assessment of biochar systems: Estimating the energetic, economic, and climate change potential. *Environ. Sci. Technol.* (2010).
- Robertson, A. D., Y. Zhang, L. A. Sherrod, S. T. Rosenzweig, L. Ma, L. Ahuja, and M. E. Schipanski, 2017a: Climate Change Impacts on Yields and Soil Carbon in Row Crop Dryland Agriculture. *J. Environ. Qual.*, doi:10.2134/jeq2017.08.0309.
- Robertson, G. P., and Coauthors, 2017b: Cellulosic biofuel contributions to a sustainable energy future: Choices and outcomes. *Science (80-.)*, **356**, eaal2324, doi:10.1126/science.aal2324. <http://www.sciencemag.org/lookup/doi/10.1126/science.aal2324> (Accessed November 11, 2018).
- , and Coauthors, 2017c: Cellulosic biofuel contributions to a sustainable energy future: Choices and outcomes. *Science (80-.)*, **356**, eaal2324, doi:10.1126/science.aal2324.
- Robledo-Abad, C., and Coauthors, 2017: Bioenergy production and sustainable development: science base for policymaking remains limited. *GCB Bioenergy*, **9**, 541–556, doi:10.1111/gcbb.12338.
- Rocha, C., 2016: Belo Horizonte: the Opportunities and Challenges of Urban Food Security Policy. http://fondazionefeltrinelli.it/app/uploads/2016/05/The-Governance-of-City-Food-Systems_The-Cases-Study-from-Around-The-World.pdf#page=29 (Accessed May 1, 2018).
- Rockström, J., M. Falkenmark, L. Karlberg, H. Hoff, S. Rost, and D. Gerten, 2009: Future water

- availability for global food production: The potential of green water for increasing resilience to global change. *Water Resour. Res.*, **45**, doi:10.1029/2007WR006767@10.1002/(ISSN)1944-7973.LANDUSE1.
<https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1029/2007WR006767%4010.1002/%28ISSN%291944-7973.LANDUSE1> (Accessed April 16, 2019).
- Roesch-McNally, G. E., S. Rabotyagov, J. C. Tyndall, G. Ettl, and S. F. Tóth, 2016: Auctioning the Forest: A Qualitative Approach to Exploring Stakeholder Responses to Bidding on Forest Ecosystem Services. *Small-scale For.*, **15**, 321–333, doi:10.1007/s11842-016-9327-0. <http://link.springer.com/10.1007/s11842-016-9327-0> (Accessed June 7, 2018).
- , A. D. Basche, J. G. Arbuckle, J. C. Tyndall, F. E. Miguez, T. Bowman, and R. Clay, 2017: The trouble with cover crops: Farmers' experiences with overcoming barriers to adoption. *Renew. Agric. Food Syst.*, 1–12, doi:10.1017/S1742170517000096. https://www.cambridge.org/core/product/identifier/S1742170517000096/type/journal_article (Accessed May 31, 2018).
- Rogelj, J., and Coauthors, 2018a: Mitigation pathways compatible with 1.5°C in the context of sustainable development. *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of 1.5 °C above pre-industrial levels and related global greenhouse gas emission pathways, in the context of strengthening the global response to the threat of climate change* <http://www.ipcc.ch/report/sr15/>.
- Rogelj, J., and Coauthors, 2018b: Scenarios towards limiting global mean temperature increase below 1.5 °C. *Nat. Clim. Chang.*, doi:10.1038/s41558-018-0091-3.
- Rogers, D., and V. Tsirkunov, 2011: *Costs and benefits of early warning systems*.
- Rojas-Downing, M. M., A. P. Nejadhashemi, T. Harrigan, and S. A. Woznicki, 2017a: Climate change and livestock: Impacts, adaptation, and mitigation. *Clim. Risk Manag.*, **16**, 145–163, doi:10.1016/j.crm.2017.02.001. <https://www.sciencedirect.com/science/article/pii/S221209631730027X> (Accessed November 8, 2018).
- , ———, ———, and ———, 2017b: Climate change and livestock: Impacts, adaptation, and mitigation. *Clim. Risk Manag.*, **16**, 145–163, doi:10.1016/j.crm.2017.02.001. <https://www.sciencedirect.com/science/article/pii/S221209631730027X> (Accessed May 30, 2018).
- Romero, H., and F. Ordenes, 2004: Emerging Urbanization in the Southern Andes. [https://doi.org/10.1659/0276-4741\(2004\)024\[0197:EUITSA\]2.0.CO;2](https://doi.org/10.1659/0276-4741(2004)024[0197:EUITSA]2.0.CO;2), **24**, 197–201, doi:10.1659/0276-4741(2004)024[0197:EUITSA]2.0.CO;2. [https://bioone.org/journals/Mountain-Research-and-Development/volume-24/issue-3/0276-4741\(2004\)024\[0197:EUITSA\]2.0.CO;2/Emerging-Urbanization-in-the-Southern-Andes/10.1659/0276-4741\(2004\)024\[0197:EUITSA\]2.0.CO;2.full](https://bioone.org/journals/Mountain-Research-and-Development/volume-24/issue-3/0276-4741(2004)024[0197:EUITSA]2.0.CO;2/Emerging-Urbanization-in-the-Southern-Andes/10.1659/0276-4741(2004)024[0197:EUITSA]2.0.CO;2.full) (Accessed April 16, 2019).
- Röös, E., B. Bajželj, P. Smith, M. Patel, D. Little, and T. Garnett, 2017: Greedy or needy? Land use and climate impacts of food in 2050 under different livestock futures. *Glob. Environ. Chang.*, **47**, 1–12, doi:10.1016/J.GLOENVCHA.2017.09.001. <https://www.sciencedirect.com/science/article/pii/S0959378016306872> (Accessed April 16, 2019).
- Rose, S. K., E. Kriegler, R. Bibas, K. Calvin, A. Popp, D. P. van Vuuren, and J. Weyant, 2014a: Bioenergy in energy transformation and climate management. *Clim. Change*, **123**, doi:10.1007/s10584-013-0965-3.
- Rose, S. K., E. Kriegler, R. Bibas, K. Calvin, A. Popp, D. P. van Vuuren, and J. Weyant, 2014b: Bioenergy in energy transformation and climate management. *Clim. Change*, **123**, 477–493, doi:10.1007/s10584-013-0965-3. <http://link.springer.com/10.1007/s10584-013-0965-3> (Accessed

November 11, 2018).

- Rosegrant, M. W., and S. A. Cline, 2003: Global food security: challenges and policies. *Science*, **302**, 1917–1919, doi:10.1126/science.1092958. <http://www.ncbi.nlm.nih.gov/pubmed/14671289> (Accessed April 16, 2019).
- Rosegrant, M. W., N. Leach, and R. V. Gerpacio, 1999: Alternative futures for world cereal and meat consumption. *Proc. Nutr. Soc.*, **58**, 219–234, doi:10.1017/S0029665199000312. http://www.journals.cambridge.org/abstract_S0029665199000312 (Accessed April 16, 2019).
- Rosenstock, T. S., K. L. Tully, C. Arias-Navarro, H. Neufeldt, K. Butterbach-Bahl, and L. V. Verchot, 2014: Agroforestry with N₂-fixing trees: Sustainable development's friend or foe? *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2013.09.001.
- Rothausen, S. G. S. A., and D. Conway, 2011: Greenhouse-gas emissions from energy use in the water sector. *Nat. Clim. Chang.*, **1**, 210–219, doi:10.1038/nclimate1147. <http://www.nature.com/articles/nclimate1147> (Accessed April 16, 2019).
- Rowe, H., and Coauthors, 2016: Integrating legacy soil phosphorus into sustainable nutrient management strategies for future food, bioenergy and water security. *Nutr. Cycl. Agroecosystems*, doi:10.1007/s10705-015-9726-1.
- Rowland, D., A. Ickowitz, B. Powell, R. Nasi, and T. Sunderland, 2017: Forest foods and healthy diets: quantifying the contributions. *Environ. Conserv.*, **44**, 102–114, doi:10.1017/S0376892916000151. https://www.cambridge.org/core/product/identifier/S0376892916000151/type/journal_article (Accessed April 5, 2019).
- Rugumamu, C., 2009: Assessment of Post-Harvest Technologies and Gender Relations in Maize Loss Reduction in Pangawe Village Eastern Tanzania. *Tanzania J. Sci.*, **35**, 67–76. <https://www.ajol.info/index.php/tjs/article/view/73533> (Accessed April 16, 2019).
- de Ruiter, H., J. Macdiarmid, L. Lynd, P. Smith, R. Matthews, and T. Kastner, 2017: *Total global agricultural land footprint associated with UK food supply 1986-2011*. Pergamon, <http://agris.fao.org/agris-search/search.do?recordID=US201700147811> (Accessed June 7, 2018).
- Rulli, M., S. Bozzi, M. Spada, D. Bocchiola, and R. Rosso, 2006: Rainfall simulations on a fire disturbed Mediterranean area. *J. Hydrol.*, **327**, 323–338. <https://www.sciencedirect.com/science/article/pii/S0022169405006189> (Accessed November 11, 2018).
- Rulli, M. C., D. Bellomi, A. Cazzoli, G. De Carolis, and P. D'Odorico, 2016: The water-land-food nexus of first-generation biofuels. *Sci. Rep.*, doi:10.1038/srep22521.
- Safriel, U., 2017: Land Degradation Neutrality (LDN) in drylands and beyond – where has it come from and where does it go. *Silva Fenn.*, **51**, 1650, doi:10.14214/sf.1650.
- Sain, G., A. M. Loboguerrero, C. Corner-Dolloff, M. Lizarazo, A. Nowak, D. Martínez-Barón, and N. Andrieu, 2017: Costs and benefits of climate-smart agriculture: The case of the Dry Corridor in Guatemala. *Agric. Syst.*, **151**, 163–173, doi:10.1016/J.AGSY.2016.05.004. <https://www.sciencedirect.com/science/article/pii/S0308521X16301160> (Accessed June 5, 2018).
- Salvati, L., A. Sabbi, D. Smiraglia, and M. Zitti, 2014a: Does forest expansion mitigate the risk of desertification? Exploring soil degradation and land-use changes in a Mediterranean country. *Int. For. Rev.*, doi:10.1505/146554814813484149.
- , ———, ———, and ———, 2014b: Does forest expansion mitigate the risk of desertification? Exploring soil degradation and land-use changes in a Mediterranean country. *Int. For. Rev.*, **16**, 485–496, doi:10.1505/146554814813484149.

- <http://www.ingentaconnect.com/content/10.1505/146554814813484149> (Accessed June 1, 2018).
- Sanchez, D. L., and D. M. Kammen, 2016a: A commercialization strategy for carbon-negative energy. *Nat. Energy*, **1**, 15002, doi:10.1038/nenergy.2015.2. <http://www.nature.com/articles/nenergy20152> (Accessed June 7, 2018).
- , and ———, 2016b: A commercialization strategy for carbon-negative energy. *Nat. Energy*, **1**, 15002, doi:10.1038/nenergy.2015.2. <http://www.nature.com/articles/nenergy20152> (Accessed April 18, 2019).
- Sánchez, J., M. D. Curt, and J. Fernández, 2017: Approach to the potential production of giant reed in surplus saline lands of Spain. *GCB Bioenergy*, doi:10.1111/gcbb.12329.
- Sanderman, J., T. Hengl, and G. J. Fiske, 2017: Soil carbon debt of 12,000 years of human land use. *Proc. Natl. Acad. Sci. U. S. A.*, **114**, 9575–9580, doi:10.1073/pnas.1706103114. <http://www.ncbi.nlm.nih.gov/pubmed/28827323> (Accessed August 13, 2018).
- Sanderson, B. M., B. C. O'Neill, and C. Tebaldi, 2016: What would it take to achieve the Paris temperature targets? *Geophys. Res. Lett.*, doi:10.1002/2016GL069563.
- Sanderson, M., and Coauthors, 2013a: Diversification and ecosystem services for conservation agriculture: Outcomes from pastures and integrated crop-livestock systems – Corrigendum. *Renew. Agric. Food Syst.*, **28**, 194, doi:10.1017/s1742170513000124. <https://www.cambridge.org/core/journals/renewable-agriculture-and-food-systems/article/diversification-and-ecosystem-services-for-conservation-agriculture-outcomes-from-pastures-and-integrated-croplivestock-systems/C29F7658818F1EC373A4F00B6E845032> (Accessed November 11, 2018).
- Sanderson, M. A., and Coauthors, 2013b: Diversification and ecosystem services for conservation agriculture: Outcomes from pastures and integrated crop-livestock systems. *Renew. Agric. Food Syst.*, **28**, 129–144, doi:10.1017/S1742170512000312. http://www.journals.cambridge.org/abstract_S1742170512000312 (Accessed May 31, 2018).
- . Santiago-Freijanes, J.J.; Rigueiro-Rodríguez, A.; Aldrey J.A.; Moreno, G.; den Herder, M.; Burgess, P. & Mosquera-Losada, M.R. 2018. [Understanding agroforestry practices in Europe through landscape features policy promotion](#). *Agroforestry systems* 92, 4. 1105-1115
- Santika, T., and Coauthors, 2017: Community forest management in Indonesia: Avoided deforestation in the context of anthropogenic and climate complexities. *Glob. Environ. Chang.*, **46**, 60–71, doi:10.1016/J.GLOENVCHA.2017.08.002. <https://www.sciencedirect.com/science/article/pii/S0959378016305933> (Accessed June 2, 2018).
- Santilli, J., 2012: *Agrobiodiversity and the Law: Regulating Genetic Resources, Food Security and Cultural Diversity*. Earthscan, New York,.
- Sapkota, T., and Coauthors, 2017: Reducing Global Warming Potential through Sustainable Intensification of Basmati Rice-Wheat Systems in India. *Sustainability*, **9**, 1044, doi:10.3390/su9061044. <http://www.mdpi.com/2071-1050/9/6/1044> (Accessed April 12, 2019).
- Sasaki, N. *et al.* Sustainable Management of Tropical Forests Can Reduce Carbon Emissions and Stabilize Timber Production. *Front. Environ. Sci.*(2016).
- Sathre, R., and J. O'Connor, 2010: Meta-analysis of greenhouse gas displacement factors of wood product substitution. *Environ. Sci. Policy*, **13**, 104–114, doi:10.1016/j.envsci.2009.12.005. <http://dx.doi.org/10.1016/j.envsci.2009.12.005>.

- Scanlon, B. R., I. Jolly, M. Sophocleous, and L. Zhang, 2007: Global impacts of conversions from natural to agricultural ecosystems on water resources: Quantity versus quality. *Water Resour. Res.*, **43**, doi:10.1029/2006WR005486. <http://doi.wiley.com/10.1029/2006WR005486> (Accessed April 15, 2019).
- Scasta, J., E. Thacker, T. Hovick, D. Engle, B. Allred, S. Fuhlendorf, and J. Weir, 2016: Patch-burn grazing (PBG) as a livestock management alternative for fire-prone ecosystems of North America. *Renew. Agric. Food Syst.*, **31**, 550–567. <https://www.cambridge.org/core/journals/renewable-agriculture-and-food-systems/article/patchburn-grazing-pbg-as-a-livestock-management-alternative-for-fireprone-ecosystems-of-north-america/FA50DEED8C4522102CD00B09B539C8F3> (Accessed November 11, 2018).
- Schatz, J., and C. J. Kucharik, 2015: Urban climate effects on extreme temperatures in Madison, Wisconsin, USA. *Environ. Res. Lett.*, **10**, 94024, doi:10.1088/1748-9326/10/9/094024. <http://stacks.iop.org/1748-9326/10/i=9/a=094024?key=crossref.296eeb5909d7fedd3e8fecf131a2a592> (Accessed June 2, 2018).
- Schjønning, P., J. L. Jensen, S. Bruun, L. S. Jensen, B. T. Christensen, L. J. Munkholm, M. Oelofse, S. Baby and L. Knudsen (2018). The Role of Soil Organic Matter for Maintaining Crop Yields: Evidence for a Renewed Conceptual Basis. *Advances in Agronomy*, 150: 35-79.
- Schipper, L., and M. Pelling, 2006: Disaster risk, climate change and international development: scope for, and challenges to, integration. *Disasters*, **30**, 19–38, doi:10.1111/j.1467-9523.2006.00304.x. <http://doi.wiley.com/10.1111/j.1467-9523.2006.00304.x> (Accessed April 19, 2019).
- Schleicher, J., 2018: The environmental and social impacts of protected areas and conservation concessions in South America. *Curr. Opin. Environ. Sustain.*, **32**, 1–8, doi:10.1016/J.COSUST.2018.01.001. <https://www.sciencedirect.com/science/article/pii/S1877343517302142> (Accessed April 4, 2019).
- Schlosser, C., K. Strzepek, and X. Gao, 2014: The Future of Global Water Stress: An Integrated Assessment. *Earth's Futur.*, **2**, 341–361, doi:10.1002/2014EF000238. Received. <http://globalchange.mit.edu/%5Cnhttp://onlinelibrary.wiley.com/doi/10.1002/2014EF000238/abstract>.
- Schmitz, O. J., and Coauthors, 2014: Animating the carbon cycle. *Ecosystems*, **17**, 344–359.
- Schmitz, O. J., C. C. Wilmers, S. J. Leroux, C. E. Doughty, T. B. Atwood, M. Galetti, A. B. Davies, and S. J. Goetz, 2018: Animals and the zoogeochemistry of the carbon cycle. *Science (80-.)*, **362**, eaar3213, doi:10.1126/SCIENCE.AAR3213. https://science.sciencemag.org/content/362/6419/eaar3213.abstract?casa_token=eqdJV2KycnIAAA:AA:3IJo_D0Qk-b_v5WaHmYYoYSG0aMWUWT7UqcUulWmKeMbKwQeBXUNOwrqK4FdOzhY-6j8vH4u4TPzFA (Accessed April 15, 2019).
- Schneider, S., and P. A. Niederle, 2010: Resistance strategies and diversification of rural livelihoods: the construction of autonomy among Brazilian family farmers. *J. Peasant Stud.*, **37**, 379–405, doi:10.1080/03066151003595168. <http://www.tandfonline.com/doi/full/10.1080/03066151003595168> (Accessed April 15, 2019).
- Schröder, P., and Coauthors, 2018: Intensify production, transform biomass to energy and novel goods and protect soils in Europe—A vision how to mobilize marginal lands. *Sci. Total Environ.*, doi:10.1016/j.scitotenv.2017.10.209.
- Schröter, D., and Coauthors, 2005: Ecosystem service supply and vulnerability to global change in

- Europe. *Science* (80-.), **310**, 1333–1337.
- Schröter, M., E. H. van der Zanden, A. P. E. van Oudenhoven, R. P. Remme, H. M. Serna-Chavez, R. S. de Groot, and P. Opdam, 2014: Ecosystem Services as a Contested Concept: a Synthesis of Critique and Counter-Arguments. *Conserv. Lett.*, **7**, 514–523, doi:10.1111/conl.12091. <http://doi.wiley.com/10.1111/conl.12091> (Accessed April 5, 2019).
- Schueler, V., S. Fuss, J. C. Steckel, U. Weddige, and T. Beringer, 2016: Productivity ranges of sustainable biomass potentials from non-agricultural land. *Environ. Res. Lett.*, doi:10.1088/1748-9326/11/7/074026.
- Schuiling, R. D., and P. Krijgsman, 2006: Enhanced Weathering: An Effective and Cheap Tool to Sequester Co₂. *Clim. Change*, **74**, 349–354, doi:10.1007/s10584-005-3485-y. <http://link.springer.com/10.1007/s10584-005-3485-y> (Accessed June 1, 2018).
- Schut, M., and Coauthors, 2016: Sustainable intensification of agricultural systems in the Central African Highlands: The need for institutional innovation. *Agric. Syst.*, **1**, 165–176. <https://www.sciencedirect.com/science/article/pii/S0308521X16300440> (Accessed May 30, 2018).
- Schwilch, G., F. Bachmann, and J. de Graaff, 2012b: Decision support for selecting SLM technologies with stakeholders. *Appl. Geogr.*, **34**, 86–98. <https://www.sciencedirect.com/science/article/pii/S0143622811002074> (Accessed November 9, 2018).
- Schwilch, G., H. P. Liniger, and H. Hurni, 2014: Sustainable Land Management (SLM) Practices in Drylands: How Do They Address Desertification Threats? *Environ. Manage.*, **54**, 983–1004, doi:10.1007/s00267-013-0071-3. <http://link.springer.com/10.1007/s00267-013-0071-3> (Accessed November 10, 2018).
- Scott, A., 2017: Making governance work for water – energy – food nexus approaches. *CKDN Work. Pap.*,. <https://www.africaportal.org/publications/making-governance-work-waterenergyfood-nexus-approaches/> (Accessed May 6, 2018).
- Scott, C. A., S. A. Pierce, M. J. Pasqualetti, A. L. Jones, B. E. Montz, and J. H. Hoover, 2011: Policy and institutional dimensions of the water–energy nexus. *Energy Policy*, **39**, 6622–6630, doi:10.1016/J.ENPOL.2011.08.013. <https://www.sciencedirect.com/science/article/pii/S0301421511006100> (Accessed April 12, 2019).
- Searchinger, T. D., T. Beringer, B. Holtzmark, D. M. Kammen, E. F. Lambin, W. Lucht, P. Raven, and J. P. van Ypersele, 2018: Europe’s renewable energy directive poised to harm global forests. *Nat. Commun.*, **9**, 10–13, doi:10.1038/s41467-018-06175-4. <http://dx.doi.org/10.1038/s41467-018-06175-4>.
- Searle, S., and C. Malins, 2014: A reassessment of global bioenergy potential in 2050. *GCB Bioenergy*, **7**, 328–336, doi:10.1111/gcbb.12141. <https://doi.org/10.1111/gcbb.12141>.
- Secretariat of the Convention on Biological Diversity, 2008: *Protected Areas in Today’s World: Their Values and Benefits for the Welfare of the Planet*.
- Seidl, R., M. J. Schelhaas, W. Rammer, and P. J. Verkerk, 2014: Increasing forest disturbances in Europe and their impact on carbon storage. *Nat. Clim. Chang.*, doi:10.1038/nclimate2318.
- Seidl, R., and Coauthors, 2017: Forest disturbances under climate change. *Nat. Clim. Chang.*, **7**, 395–402, doi:10.1038/nclimate3303
- Seinfeld, J. H., and S. N. Pandis, *Atmospheric chemistry and physics: from air pollution to climate change*. <https://www.wiley.com/en-us/Atmospheric+Chemistry+and+Physics%3A+From+Air+Pollution+to+Climate+Change%2C+3rd>

+Edition-p-9781118947401 (Accessed June 7, 2018).

- Sen, A., 1992: Inequality Reexamined Mass. https://scholar.google.com/scholar?hl=en&as_sdt=0%2C5&q=Sen%2C+A.+K.+%281992%29.+Inequality+Reexamined.+Oxford%3A+Oxford+University+Press.&btnG= (Accessed November 10, 2018).
- Seo, S. N., 2010: Is an integrated farm more resilient against climate change? A micro-econometric analysis of portfolio diversification in African agriculture. *Food Policy*, **35**, 32–40, doi:10.1016/J.FOODPOL.2009.06.004. <https://www.sciencedirect.com/science/article/pii/S030691920900058X> (Accessed April 4, 2019).
- Seppelt, R., C. F. Dormann, F. V. Eppink, S. Lautenbach, and S. Schmidt, 2011: A quantitative review of ecosystem service studies: approaches, shortcomings and the road ahead. *J. Appl. Ecol.*, **48**, 630–636, doi:10.1111/j.1365-2664.2010.01952.x. <http://doi.wiley.com/10.1111/j.1365-2664.2010.01952.x> (Accessed April 5, 2019).
- Shackley, S., J. Hammond, J. Gaunt, and R. Ibarrola, 2011a: The feasibility and costs of biochar deployment in the UK. *Carbon Manag.*, **2**, 335–356, doi:10.4155/cmt.11.22. <http://www.tandfonline.com/doi/abs/10.4155/cmt.11.22> (Accessed May 30, 2018).
- , ———, ———, and ———, 2011b: The feasibility and costs of biochar deployment in the UK. *Carbon Manag.*, **2**, 335–356, doi:10.4155/cmt.11.22. <http://www.tandfonline.com/doi/abs/10.4155/cmt.11.22> (Accessed April 18, 2019).
- Shah, T., M. Giordano, and A. Mukherji, 2012: Political economy of the energy-groundwater nexus in India: exploring issues and assessing policy options. *Hydrogeol. J.*, **20**, 995–1006, doi:10.1007/s10040-011-0816-0. <http://link.springer.com/10.1007/s10040-011-0816-0> (Accessed June 2, 2018).
- Shaw, C., S. Hales, P. Howden-Chapman, and R. Edwards, 2014: Health co-benefits of climate change mitigation policies in the transport sector. *Nat. Clim. Chang.*, **4**, 427–433, doi:10.1038/nclimate2247. <http://www.nature.com/articles/nclimate2247> (Accessed June 7, 2018).
- Shcherbak, I., N. Millar, and G. P. Robertson, 2014: Global metaanalysis of the nonlinear response of soil nitrous oxide (N₂O) emissions to fertilizer nitrogen. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 9199–9204, doi:10.1073/pnas.1322434111. <http://www.ncbi.nlm.nih.gov/pubmed/24927583> (Accessed June 3, 2018).
- Sheahan, M., and C. B. Barrett, 2017a: Ten striking facts about agricultural input use in Sub-Saharan Africa. *Food Policy*, **67**, 12–25, doi:10.1016/J.FOODPOL.2016.09.010. <https://www.sciencedirect.com/science/article/pii/S0306919216303773> (Accessed June 7, 2018).
- , and C. B. Barrett, 2017b: Food loss and waste in Sub-Saharan Africa: A critical review. *Food Policy*, **70**, 1–12, doi:10.1016/j.foodpol.2017.03.012. <https://www.sciencedirect.com/science/article/pii/S0306919217302440> (Accessed May 10, 2018).
- Shen, X., L. Wang, C. Wu, T. Lv, Z. Lu, W. Luo, and G. Li, 2017: Local interests or centralized targets? How China’s local government implements the farmland policy of Requisition–Compensation Balance. *Land use policy*, **67**, 716–724, doi:10.1016/J.LANDUSEPOL.2017.06.012. <https://www.sciencedirect.com/science/article/pii/S0264837715301952> (Accessed May 31, 2018).
- Shindell, D., and Coauthors, 2012: Simultaneously mitigating near-term climate change and improving human health and food security. *Science (80-.)*, **335**, 183–189. <http://science.sciencemag.org/content/335/6065/183.short> (Accessed November 12, 2018).
- Shisanya, S., and P. Mafongoya, 2016: Adaptation to climate change and the impacts on household food security among rural farmers in uMzinyathi District of Kwazulu-Natal, South Africa. *Food Secur.*,

- 8, 597–608, doi:10.1007/s12571-016-0569-7. <http://link.springer.com/10.1007/s12571-016-0569-7> (Accessed June 1, 2018).
- Shively, G., and G. Thapa, 2016: Markets, transportation infrastructure, and food prices in Nepal. *Am. J. Agric. Econ.*, **99**, aaw086, doi:10.1093/ajae/aaw086. <https://academic.oup.com/ajae/article-lookup/doi/10.1093/ajae/aaw086> (Accessed April 15, 2019).
- Sida, T. S., F. Baudron, K. Hadgu, A. Derero, and K. E. Giller, 2018: Crop vs. tree: Can agronomic management reduce trade-offs in tree-crop interactions? *Agric. Ecosyst. Environ.*, **260**, 36–46, doi:10.1016/J.AGEE.2018.03.011. <https://www.sciencedirect.com/science/article/pii/S0167880918301269> (Accessed April 5, 2019).
- Sikkema, R., M. Junginger, P. McFarlane, and A. Faaij, 2013: The GHG contribution of the cascaded use of harvested wood products in comparison with the use of wood for energy-A case study on available forest resources in Canada. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2013.03.007.
- Silalahi, M., A. Utomo, T. Walsh, A. Ayat, Andriansyah, and S. Bashir, 2017: Indonesia's ecosystem restoration concessions. *Unasylva*, **68**, 63–70.
- Sileshi, G., E. Kuntashula, P. Matakala, P. N.-A. Systems, and U. 2008, 2008: Farmers' perceptions of tree mortality, pests and pest management practices in agroforestry in Malawi, Mozambique and Zambia. *Springer*,. <https://link.springer.com/article/10.1007/s10457-007-9082-5> (Accessed May 30, 2018).
- Silva-Olaya, A. M., C. E. P. Cerri, S. Williams, C. C. Cerri, C. A. Davies, and K. Paustian, 2017: Modelling SOC response to land use change and management practices in sugarcane cultivation in South-Central Brazil. *Plant Soil*, doi:10.1007/s11104-016-3030-y.
- da Silva, J., S. Kernaghan, and A. Luque, 2012: A systems approach to meeting the challenges of urban climate change. *Int. J. Urban Sustain. Dev.*, **4**, 125–145, doi:10.1080/19463138.2012.718279. <http://www.tandfonline.com/doi/abs/10.1080/19463138.2012.718279> (Accessed April 4, 2019).
- Sims, R., and Coauthors, 2014: Transport. *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- Singh, G., 2009: Salinity-related desertification and management strategies: Indian experience. *L. Degrad. Dev.*, **20**, 367–385, doi:10.1002/ldr.933. <http://doi.wiley.com/10.1002/ldr.933> (Accessed November 12, 2018).
- Singh, S. N., and A. Verma, 2007a: Environmental Review: The Potential of Nitrification Inhibitors to Manage the Pollution Effect of Nitrogen Fertilizers in Agricultural and Other Soils: A Review. *Environ. Pract.*, **9**, 266–279, doi:10.1017/S1466046607070482. <https://www.tandfonline.com/doi/full/10.1017/S1466046607070482> (Accessed June 7, 2018).
- , and —, 2007b: Environmental Review: The Potential of Nitrification Inhibitors to Manage the Pollution Effect of Nitrogen Fertilizers in Agricultural and Other Soils: A Review. *Environ. Pract.*, **9**, 266–279, doi:10.1017/S1466046607070482. <https://www.tandfonline.com/doi/full/10.1017/S1466046607070482> (Accessed April 18, 2019).
- Skees, J. R., and B. Collier, 2012: The Roles of Weather Insurance and the Carbon Market. *Greening the Financial Sector*, Springer Berlin Heidelberg, Berlin, Heidelberg, 111–164 http://link.springer.com/10.1007/978-3-642-05087-9_4 (Accessed June 1, 2018).
- Slade, R., A. Bauen, and R. Gross, 2014: Global bioenergy resources. *Nat. Clim. Chang.*, doi:10.1038/nclimate2097.

- Smit, B., and Coauthors, 2001: Adaptation to Climate Change in the Context of Sustainable Development and Equity. *Climate Change 2001: Impacts, Adaptation and Vulnerability. Contribution of Working Group II to the Third Assessment Report of the Intergovernmental Panel on Climate Change*.
- Smit, J., and J. Nasr, 1992: Urban agriculture for sustainable cities: using wastes and idle land and water bodies as resources. *Environ. Urban.*, **4**, 141–152, doi:10.1177/095624789200400214. <http://journals.sagepub.com/doi/10.1177/095624789200400214> (Accessed April 15, 2019).
- Smit, W., 2016: Urban governance and urban food systems in Africa: Examining the linkages. *Cities*, <https://www.sciencedirect.com/science/article/pii/S026427511630083X> (Accessed May 1, 2018).
- Smith, D. W., 1998: Urban Food Systems and the Poor in Developing Countries. *Trans. Inst. Br. Geogr.*, **23**, 207–219, doi:10.1111/j.0020-2754.1998.00207.x. <http://doi.wiley.com/10.1111/j.0020-2754.1998.00207.x> (Accessed April 15, 2019).
- Smith, P., 2006: Monitoring and verification of soil carbon changes under Article 3.4 of the Kyoto Protocol. *Soil Use Manag.*, **20**, 264–270, doi:10.1111/j.1475-2743.2004.tb00367.x. <http://doi.wiley.com/10.1111/j.1475-2743.2004.tb00367.x> (Accessed May 30, 2018).
- Smith, P., 2008: Land use change and soil organic carbon dynamics. *Nutr. Cycl. Agroecosystems*, **81**, 169–178, doi:10.1007/s10705-007-9138-y. <http://link.springer.com/10.1007/s10705-007-9138-y> (Accessed November 7, 2018).
- , 2013: Delivering food security without increasing pressure on land. *Glob. Food Sec.*, **2**, 18–23, doi:10.1016/j.gfs.2012.11.008. <https://www.sciencedirect.com/science/article/pii/S2211912412000363> (Accessed November 10, 2018).
- , 2016a: Soil carbon sequestration and biochar as negative emission technologies. *Glob. Chang. Biol.*, **22**, 1315–1324, doi:10.1111/gcb.13178. <http://doi.wiley.com/10.1111/gcb.13178> (Accessed November 10, 2018).
- , and P. J. Gregory, 2013: Climate change and sustainable food production. *Proc. Nutr. Soc.*, **72**, 21–28, doi:10.1017/S0029665112002832. http://www.journals.cambridge.org/abstract_S0029665112002832 (Accessed April 15, 2019).
- Smith, P., and Coauthors, Greenhouse gas mitigation in agriculture. *rstb.royalsocietypublishing.org*.
- Smith, P., S. Davis, F. Creutzig, S. Fuss, ... J. M.-N. C., and undefined 2016, Biophysical and economic limits to negative CO₂ emissions. *nature.com*, <https://www.nature.com/articles/nclimate2870> (Accessed November 9, 2018b).
- Smith, P., and Coauthors, 2007: *Agriculture. In Climate Change 2007: Mitigation. Contribution of Working Group III to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change*. B. Metz, O. Davidson, P. Bosch, R. Dave, and L. Meyer, Eds. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.,
- , and Coauthors, 2008: Greenhouse gas mitigation in agriculture. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, **363**, 789–813, doi:10.1098/rstb.2007.2184. <http://rstb.royalsocietypublishing.org/content/363/1492/789.short> (Accessed May 30, 2018).
- , and Coauthors, 2013: How much land-based greenhouse gas mitigation can be achieved without compromising food security and environmental goals? *Glob. Chang. Biol.*, **19**, 2285–2302, doi:10.1111/gcb.12160. <http://doi.wiley.com/10.1111/gcb.12160> (Accessed November 11, 2018).
- Smith, P., and Coauthors, 2014a: Agriculture, Forestry and Other Land Use (AFOLU). *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, O. Edenhofer et al., Eds., Cambridge

- University Press, Cambridge, United Kingdom and New York, NY, USA <http://pure.iiasa.ac.at/11115/> (Accessed December 20, 2017).
- , and Coauthors, 2014b: Agriculture, Forestry and Other Land Use (AFOLU). *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* <http://pure.iiasa.ac.at/11115/> (Accessed December 20, 2017).
- Smith, P., and Coauthors, 2015: Biogeochemical cycles and biodiversity as key drivers of ecosystem services provided by soils. **1**, 665–685, doi:10.5194/soil-1-665-2015. www.soil-journal.net/1/665/2015/ (Accessed May 30, 2018).
- , and Coauthors, 2016a: Biophysical and economic limits to negative CO₂ emissions. *Nat. Clim. Chang.*, **6**, 42–50, doi:10.1038/NCLIMATE2870. <https://www.nature.com/articles/nclimate2870> (Accessed May 1, 2018).
- , and Coauthors, 2016b: Global change pressures on soils from land use and management. *Glob. Chang. Biol.*, **22**, 1008–1028, doi:10.1111/gcb.13068. <http://doi.wiley.com/10.1111/gcb.13068> (Accessed April 4, 2019).
- , and Coauthors, 2016c: Biophysical and economic limits to negative CO₂ emissions. *Nat. Clim. Chang.*, **6**, 42–50. <https://www.nature.com/articles/nclimate2870> (Accessed May 1, 2018).
- , R. S. Haszeldine, and S. M. Smith, 2016d: Preliminary assessment of the potential for, and limitations to, terrestrial negative emission technologies in the UK. *Environ. Sci. Process. Impacts*, doi:10.1039/C6EM00386A.
- Smith, S. V., W. H. Renwick, R. W. Buddemeier, and C. J. Crossland, 2001: Budgets of soil erosion and deposition for sediments and sedimentary organic carbon across the conterminous United States. *Global Biogeochem. Cycles*, **15**, 697–707, doi:10.1029/2000GB001341. <http://doi.wiley.com/10.1029/2000GB001341> (Accessed November 11, 2018).
- , R. O. Slezzer, W. H. Renwick, and R. W. Buddemeier, 2005: FATES OF ERODED SOIL ORGANIC CARBON: MISSISSIPPI BASIN CASE STUDY. *Ecol. Appl.*, **15**, 1929–1940, doi:10.1890/05-0073. <http://doi.wiley.com/10.1890/05-0073> (Accessed November 11, 2018).
- Smith, V. H., and B. K. Goodwin, 1996: Crop Insurance, Moral Hazard, and Agricultural Chemical Use. *Am. J. Agric. Econ.*, **78**, 428, doi:10.2307/1243714. <https://academic.oup.com/ajae/article-lookup/doi/10.2307/1243714> (Accessed April 15, 2019).
- Smith, V. H., and J. W. Glauber, 2012: Agricultural Insurance in Developed Countries: Where Have We Been and Where Are We Going? *Appl. Econ. Perspect. Policy*, **34**, 363–390, doi:10.1093/aep/pps029. <https://academic.oup.com/aep/article-lookup/doi/10.1093/aep/pps029> (Accessed May 31, 2018).
- Smyth, C. E., G. Stinson, E. Neilson, T. C. Lemprière, M. Hafer, G. J. Rampley, and W. A. Kurz, 2014: Quantifying the biophysical climate change mitigation potential of Canada’s forest sector. *Biogeosciences*, **11**, 3515–3529, doi:10.5194/bg-11-3515-2014.
- SMYTH, C., G. RAMPLEY, T. C. LEMPRI, O. SCHWAB, and W. A. KURZ, 2016: Estimating product and energy substitution benefits in national-scale mitigation analyses for Canada. *Glob. Chang. Biol. Bioenergy*, 1–14, doi:10.1111/gcbb.12389.
- Soane, B. D., and C. van Ouwerkerk, 1994: Soil Compaction Problems in World Agriculture. *Dev. Agric. Eng.*, **11**, 1–21, doi:10.1016/B978-0-444-88286-8.50009-X. <https://www.sciencedirect.com/science/article/pii/B978044488286850009X> (Accessed May 31, 2018).

- Sohi, S., 2012: Carbon Storage with Benefits. *Sci.*, **338**, 1034–1035, doi:10.1126/science.1227620. <http://www.ncbi.nlm.nih.gov/pubmed/22997311> (Accessed June 7, 2018).
- Sommer, R. & Bossio, D. Dynamics and climate change mitigation potential of soil organic carbon sequestration. *J. Environ. Manage.* (2014).
- Song, G., M. Li, P. Fullana-i-Palmer, D. Williamson, and Y. Wang, 2017: Dietary changes to mitigate climate change and benefit public health in China. *Sci. Total Environ.*, **577**, 289–298, doi:10.1016/j.scitotenv.2016.10.184. <https://www.sciencedirect.com/science/article/pii/S0048969716323725> (Accessed November 8, 2018).
- Song, W., and X. Deng, 2015: Effects of Urbanization-Induced Cultivated Land Loss on Ecosystem Services in the North China Plain. *Energies*, **8**, 5678–5693, doi:10.3390/en8065678. <http://www.mdpi.com/1996-1073/8/6/5678> (Accessed May 31, 2018).
- Sonntag, S., J. Pongratz, C. H. Reick, and H. Schmidt, 2016: Reforestation in a high-CO₂ world—Higher mitigation potential than expected, lower adaptation potential than hoped for. *Geophys. Res. Lett.*, doi:10.1002/2016GL068824.
- Soussana, J., and Coauthors, 2019: Matching policy and science: Rationale for the “4 per 1000-soils for food security and climate” initiative. *Soil Tillage Res.*, **188**, 3–15. <https://www.sciencedirect.com/science/article/pii/S0167198717302271> (Accessed November 11, 2018).
- Specht, K., R. Siebert, I. Hartmann, and ... U. F., 2014: Urban agriculture of the future: an overview of sustainability aspects of food production in and on buildings. *Agric. Human Values.*, <https://link.springer.com/article/10.1007/s10460-013-9448-4> (Accessed May 1, 2018).
- Springer, N. P., and Coauthors, 2015: Sustainable Sourcing of Global Agricultural Raw Materials: Assessing Gaps in Key Impact and Vulnerability Issues and Indicators. *PLoS One*, **10**, e0128752, doi:10.1371/journal.pone.0128752. <https://dx.plos.org/10.1371/journal.pone.0128752> (Accessed April 19, 2019).
- Springmann, M., and Coauthors, 2016: Global and regional health effects of future food production under climate change: a modelling study. *Lancet*, **387**, 1937–1946, doi:10.1016/S0140-6736(15)01156-3. <https://www.sciencedirect.com/science/article/pii/S0140673615011563> (Accessed April 4, 2019).
- , and Coauthors, 2018a: Options for keeping the food system within environmental limits. *Nature*, **562**, 519–525, doi:10.1038/s41586-018-0594-0. <http://www.nature.com/articles/s41586-018-0594-0> (Accessed November 11, 2018).
- , and Coauthors, 2018b: Options for keeping the food system within environmental limits. *Nature*, **562**, 519–525, doi:10.1038/s41586-018-0594-0. <http://www.nature.com/articles/s41586-018-0594-0> (Accessed November 11, 2018).
- , and Coauthors, 2018c: Options for keeping the food system within environmental limits. *Nature*, doi:10.1038/s41586-018-0594-0.
- Squires, V., and E. Karami, 2005: Livestock management in the Arid zone: Coping strategies. *J. Rangel. Sci.*, **5**, 336–346. http://www.rangeland.ir/article_520478.html (Accessed November 11, 2018).
- Srinivasa Rao, C., K. A. Gopinath, J. V. N. S. Prasad, Prasannakumar, and A. K. Singh, 2016: Climate Resilient Villages for Sustainable Food Security in Tropical India: Concept, Process, Technologies, Institutions, and Impacts. *Adv. Agron.*, **140**, 101–214, doi:10.1016/BS.AGRON.2016.06.003. <https://www.sciencedirect.com/science/article/pii/S0065211316300797> (Accessed April 13, 2019).
- Srinivasan, V., S. Thompson, K. Madhyastha, G. Penny, K. Jeremiah, and S. Lele, 2015: Why is the

- Arkavathy River drying? A multiple-hypothesis approach in a data-scarce region. *Hydrol. Earth Syst. Sci.*, **19**, 1950–2015, doi:10.5194/hess-19-1905-2015. www.hydrol-earth-syst-sci.net/19/1905/2015/ (Accessed June 2, 2018).
- Stallard, R. F., 1998: Terrestrial sedimentation and the carbon cycle: Coupling weathering and erosion to carbon burial. *Global Biogeochem. Cycles*, **12**, 231–257, doi:10.1029/98GB00741. <http://doi.wiley.com/10.1029/98GB00741> (Accessed November 11, 2018).
- Stanturf, J. A., P. Kant, J.-P. B. Lillesø, S. Mansourian, M. Kleine, L. Graudal, and P. Madsen, 2015: *Forest Landscape Restoration as a Key Component of Climate Change Mitigation and Adaptation*. Vienna, Austria, http://curis.ku.dk/ws/files/161428268/Stanturff_et_al_2015_IUFRO_World_Series_vol_34_FLR_adaptation_mitigation.pdf (Accessed June 7, 2018).
- Stathers, T., R. Lamboll, and B. M. Mvumi, 2013: Postharvest agriculture in changing climates: its importance to African smallholder farmers. *Food Secur.*, **5**, 361–392, doi:10.1007/s12571-013-0262-z.
- Stefan, V., E. van Herpen, A. A. Tudoran, and L. Lähteenmäki, 2013: Avoiding food waste by Romanian consumers: The importance of planning and shopping routines. *Food Qual. Prefer.*, **28**, 375–381, doi:10.1016/J.FOODQUAL.2012.11.001. <https://www.sciencedirect.com/science/article/abs/pii/S0950329312002066> (Accessed April 15, 2019).
- Stehfest, E., L. Bouwman, D. P. Van Vuuren, M. G. J. Den Elzen, B. Eickhout, and P. Kabat, 2009: Climate benefits of changing diet. *Clim. Change*, **95**, 83–102, doi:10.1007/s10584-008-9534-6.
- Steinbach, H., R. A.-J. of E. Quality, and undefined 2006, Changes in soil organic carbon contents and nitrous oxide emissions after introduction of no-till in Pampean agroecosystems. *dl.sciencesocieties.org*,. <https://dl.sciencesocieties.org/publications/jeq/abstracts/35/1/3> (Accessed November 11, 2018).
- Stellmes, M., A. Röder, T. Udelhoven, and J. Hill, 2013: Mapping syndromes of land change in Spain with remote sensing time series, demographic and climatic data. *Land use policy*, **30**, 685–702, doi:10.1016/j.landusepol.2012.05.007. <http://linkinghub.elsevier.com/retrieve/pii/S0264837712000920> (Accessed November 11, 2018).
- Sterling, S. M., A. Ducharne, and J. Polcher, 2013: The impact of global land-cover change on the terrestrial water cycle. *Nat. Clim. Chang.*, doi:10.1038/Nclimate1690.
- Sternberg, T., and B. Batbuyan, 2013: Integrating the Hyogo Framework into Mongolia’s disaster risk reduction (DRR) policy and management. *Int. J. Disaster Risk Reduct.*, doi:10.1016/j.ijdr.2013.05.003.
- Stevanovic, M., and Coauthors, 2016: The impact of high-end climate change on agricultural welfare. *Sci. Adv.*, **2**, e1501452–e1501452, doi:10.1126/sciadv.1501452. <http://advances.sciencemag.org/cgi/doi/10.1126/sciadv.1501452> (Accessed June 7, 2018).
- Stevanović, M., and Coauthors, 2017: Mitigation Strategies for Greenhouse Gas Emissions from Agriculture and Land-Use Change: Consequences for Food Prices. *Environ. Sci. Technol.*, **51**, 365–374, doi:10.1021/acs.est.6b04291. <http://pubs.acs.org/doi/10.1021/acs.est.6b04291> (Accessed April 14, 2019).
- Steward, A., 2007: Nobody farms here anymore: Livelihood diversification in the Amazonian community of Carvão, a historical perspective. *Agric. Human Values*, **24**, 75–92, doi:10.1007/s10460-006-9032-2. <http://link.springer.com/10.1007/s10460-006-9032-2> (Accessed April 15, 2019).
- Stibig, H.-J., F. Achard, S. Carboni, R. Raši, and J. Miettinen, 2014: Change in tropical forest cover of

- Southeast Asia from 1990 to 2010. *Biogeosciences*, **11**, 247–258, doi:10.5194/bg-11-247-2014. www.biogeosciences.net/11/247/2014/ (Accessed June 2, 2018).
- Stigter, C. ., M. V. . Sivakumar, and D. . Rijks, 2000: Agrometeorology in the 21st century: workshop summary and recommendations on needs and perspectives. *Agric. For. Meteorol.*, **103**, 209–227, doi:10.1016/S0168-1923(00)00113-1. <https://www.sciencedirect.com/science/article/pii/S0168192300001131> (Accessed April 15, 2019).
- Stoll-Kleemann, S., and T. O’Riordan, 2015: The Sustainability Challenges of Our Meat and Dairy Diets. *Environ. Sci. Policy Sustain. Dev.*, **57**, 34–48, doi:10.1080/00139157.2015.1025644. <http://www.tandfonline.com/doi/full/10.1080/00139157.2015.1025644> (Accessed April 15, 2019).
- Stone, B., J. J. Hess, and H. Frumkin, 2010: Urban Form and Extreme Heat Events: Are Sprawling Cities More Vulnerable to Climate Change Than Compact Cities? *Environ. Health Perspect.*, **118**, 1425–1428, doi:10.1289/ehp.0901879. <https://ehp.niehs.nih.gov/doi/10.1289/ehp.0901879> (Accessed November 11, 2018).
- Strassburg, B. B. N., A. E. Latawiec, L. G. Barioni, C. A. Nobre, V. P. da Silva, J. F. Valentim, M. Vianna, and E. D. Assad, 2014: When enough should be enough: Improving the use of current agricultural lands could meet production demands and spare natural habitats in Brazil. *Glob. Environ. Chang.*, **28**, 84–97, doi:10.1016/J.GLOENVCHA.2014.06.001. <https://www.sciencedirect.com/science/article/pii/S0959378014001046> (Accessed June 2, 2018).
- Strengers, B. J., J. G. Van Minnen, and B. Eickhout, 2008: The role of carbon plantations in mitigating climate change: potentials and costs. *Clim. Change*, **88**, 343–366, doi:10.1007/s10584-007-9334-4. <http://link.springer.com/10.1007/s10584-007-9334-4> (Accessed June 1, 2018).
- Stringer, L., J. Dyer, M. Reed, A. Dougill, C. Twyman, and D. Mkwambisi, 2009: Adaptations to climate change, drought and desertification: local insights to enhance policy in southern Africa. *Environ. Sci. Policy*, **12**, 748–765. <https://www.sciencedirect.com/science/article/pii/S1462901109000604> (Accessed November 11, 2018).
- Stringer, L. C., A. J. Dougill, D. D. Mkwambisi, J. C. Dyer, F. K. Kalaba, and M. Mngoli, 2012: Challenges and opportunities for carbon management in Malawi and Zambia. *Carbon Manag.*, **3**, 159–173, doi:10.4155/cmt.12.14. <http://www.tandfonline.com/doi/abs/10.4155/cmt.12.14> (Accessed November 10, 2018).
- Suckall, N., L. C. Stringer, and E. L. Tompkins, 2015: Presenting Triple-Wins? Assessing Projects That Deliver Adaptation, Mitigation and Development Co-benefits in Rural Sub-Saharan Africa. *Ambio*, **44**, 34–41, doi:10.1007/s13280-014-0520-0. <http://link.springer.com/10.1007/s13280-014-0520-0> (Accessed April 15, 2019).
- Suding, K., and Coauthors, 2015: Committing to ecological restoration. *Science (80-.)*, **348**, 638 LP-640.
- Sunderland, T. C. H., B. Powell, A. Ickowitz, S. Foli, M. Pinedo-Vasquez, R. Nasi, and C. Padoch, 2013: Food security and nutrition: The role of forests. <https://www.cifor.org/library/4103/> (Accessed April 5, 2019).
- Swanton, C. J., S. D. Murphy, D. J. Hume, and D. Clements, 1996: Recent improvements in the energy efficiency of agriculture: case studies from Ontario, Canada. *Agric. Syst.*, **52**, 399–418.
- Taboada, M., G. Rubio, and E. Chaneton, 2011: Grazing impacts on soil physical, chemical, and ecological properties in forage production systems. *Soil management: building a stable base for agriculture*, 301–320 http://www.academia.edu/download/41429168/Grazing_Impacts_on_Soil_Physical_Chemica20160122-3852-11t0a0f.pdf (Accessed November 12, 2018).
- Tacconi, L., 2016: Preventing fires and haze in Southeast Asia. *Nat. Clim. Chang.*, **6**, 640–643,

- doi:10.1038/nclimate3008. <http://www.nature.com/articles/nclimate3008> (Accessed June 7, 2018).
- Tadasse, G., B. Algieri, M. Kalkuhl, and J. von Braun, 2016: Drivers and Triggers of International Food Price Spikes and Volatility. *Food Price Volatility and Its Implications for Food Security and Policy*, Springer International Publishing, Cham, 59–82 http://link.springer.com/10.1007/978-3-319-28201-5_3 (Accessed June 1, 2018).
- Tai, A. P. K., M. V. Martin, and C. L. Heald, 2014: Threat to future global food security from climate change and ozone air pollution. *Nat. Clim. Chang.*, **4**, 817–821, doi:10.1038/nclimate2317. <http://www.nature.com/articles/nclimate2317> (Accessed June 7, 2018).
- Tan, R., V. Beckmann, L. van den Berg, and F. Qu, 2009: Governing farmland conversion: Comparing China with the Netherlands and Germany. *Land use policy*, **26**, 961–974, doi:10.1016/J.LANDUSEPOL.2008.11.009. <https://www.sciencedirect.com/science/article/pii/S0264837708001543> (Accessed May 31, 2018).
- Tansey, K., and Coauthors, 2004: Vegetation burning in the year 2000: Global burned area estimates from SPOT VEGETATION data. *J. Geophys. Res. Atmos.*, **109**, D14. https://scholar.google.com/scholar?hl=en&as_sdt=0%2C5&q=Tansey+K.%2C+Grégoire+J.M.%2C+Stroppiana+D.%2C+Sousa+A.%2C+Silva+J.%2C+Pereira+J.M.C.%2C+Boschetti+L.%2C+Maggi+M.%2C+Brivio+P.A.%2C+Fraser+R.%2C+Flasse+S.%2C+Ershov+D.%2C+Binaghi+E.%2C+Gratz+D.%2C+P (Accessed November 12, 2018).
- Tao, Y., F. Li, X. Liu, D. Zhao, X. Sun, and L. X. Modelling, 2015: Variation in ecosystem services across an urbanization gradient: A study of terrestrial carbon stocks from Changzhou, China. *Ecol. Appl.*, <https://www.sciencedirect.com/science/article/pii/S0304380015001994> (Accessed May 1, 2018).
- Tarnocai, C., 2006: The effect of climate change on carbon in Canadian peatlands. *Glob. Planet. Change*, **53**, 222–232, doi:10.1016/J.GLOPLACHA.2006.03.012. <https://www.sciencedirect.com/science/article/pii/S0921818106001202> (Accessed April 4, 2019).
- Taylor, C., and Coauthors, 2017: Global Coverage of Agricultural Sustainability Standards, and Their Role in Conserving Biodiversity. *Conserv. Lett.*, **10**, 610–618, doi:10.1111/conl.12314. <http://doi.wiley.com/10.1111/conl.12314> (Accessed April 19, 2019).
- Taylor, L. L., and Coauthors, 2016: Enhanced weathering strategies for stabilizing climate and averting ocean acidification. *Nat. Clim. Chang.*, **6**, 402–406, doi:10.1038/nclimate2882. <http://www.nature.com/articles/nclimate2882> (Accessed November 11, 2018).
- TEEB, 2009: *TEEB For Policy Makers: Responding To The Value Of Nature. The Economics of Ecosystems and Biodiversity Project*.
- Tefera, T., 2012: Post-harvest losses in African maize in the face of increasing food shortage. *Food Secur.*, **4**, 267–277, doi:10.1007/s12571-012-0182-3. <http://link.springer.com/10.1007/s12571-012-0182-3> (Accessed April 15, 2019).
- Temba, B. A., Y. Sultanbawa, D. J. Kriticos, G. P. Fox, J. J. W. Harvey, and M. T. Fletcher, 2016: Tools for Defusing a Major Global Food and Feed Safety Risk: Nonbiological Postharvest Procedures to Decontaminate Mycotoxins in Foods and Feeds. *J. Agric. Food Chem.*, doi:10.1021/acs.jafc.6b03777.
- The World Bank, 2011: *Rising Global Interest in Farmland: Can It Yield Sustainable and Equitable Benefits?* World Bank, Washington DC, <http://elibrary.worldbank.org/doi/book/10.1596/978-0-8213-8591-3>.
- Thomaier, S., K. Specht, D. Henckel, A. Dierich, R. Siebert, U. B. Freisinger, and M. Sawicka, 2015: Farming in and on urban buildings: Present practice and specific novelties of Zero-Acreage Farming

- (ZFarming). *Renew. Agric. Food Syst.*, **30**, 43–54, doi:10.1017/S1742170514000143. http://www.journals.cambridge.org/abstract_S1742170514000143 (Accessed June 2, 2018).
- Thomalla, F., T. Downing, E. Spanger-Siegfried, G. Han, and J. Rockström, 2006: Reducing hazard vulnerability: Towards a common approach between disaster risk reduction and climate adaptation. *Disasters*, doi:10.1111/j.1467-9523.2006.00305.x.
- Thormark, C., 2006: The effect of material choice on the total energy need and recycling potential of a building. *Build. Environ.*, **41**, 1019–1026, doi:10.1016/J.BUILDENV.2005.04.026. <https://www.sciencedirect.com/science/article/pii/S0360132305001605> (Accessed April 15, 2019).
- Thornbush, M., O. Golubchikov, and S. Bouzarovski, 2013: Sustainable cities targeted by combined mitigation–adaptation efforts for future-proofing. *Sustain. Cities Soc.*, **9**, 1–9, doi:10.1016/j.scs.2013.01.003. <https://linkinghub.elsevier.com/retrieve/pii/S2210670713000048> (Accessed November 11, 2018).
- Thornton, P., J. van de Steeg, A. Notenbaert, and M. Herrero, 2009a: The impacts of climate change on livestock and livestock systems in developing countries: A review of what we know and what we need to know. *Agric. Syst.*, **101**, 113–127. <https://www.sciencedirect.com/science/article/pii/S0308521X09000584> (Accessed May 30, 2018).
- Thornton, P. E., and Coauthors, 2017: Biospheric feedback effects in a synchronously coupled model of human and Earth systems. *Nat. Clim. Chang.*, **7**, 496–500, doi:10.1038/nclimate3310. <http://www.nature.com/articles/nclimate3310> (Accessed April 14, 2019).
- Thornton, P. K., and M. Herrero, 2014: Climate change adaptation in mixed crop–livestock systems in developing countries. *Glob. Food Sec.*, **3**, 99–107, doi:10.1016/J.GFS.2014.02.002. <https://www.sciencedirect.com/science/article/pii/S2211912414000108> (Accessed June 1, 2018).
- Thornton, P. K., J. van de Steeg, A. Notenbaert, and M. Herrero, 2009b: The impacts of climate change on livestock and livestock systems in developing countries: A review of what we know and what we need to know. *Agric. Syst.*, **101**, 113–127, doi:10.1016/J.AGSY.2009.05.002. <https://www.sciencedirect.com/science/article/pii/S0308521X09000584> (Accessed April 4, 2019).
- Thyberg, K. L., and D. J. Tonjes, 2016: Drivers of food waste and their implications for sustainable policy development. *Resour. Conserv. Recycl.*, **106**, 110–123, doi:10.1016/J.RESCONREC.2015.11.016. <https://www.sciencedirect.com/science/article/pii/S0921344915301439> (Accessed April 15, 2019).
- Tian, X., B. Sohngen, J. Baker, S. Ohrel, and A. A. Fawcett, 2018: Will U.S. Forests Continue to Be a Carbon Sink? *Land Econ.*, **94**, 97–113, doi:10.3368/le.94.1.97.
- Tigchelaar, M., D. S. Battisti, R. L. Naylor, and D. K. Ray, 2018: Future warming increases probability of globally synchronized maize production shocks. *Proc. Natl. Acad. Sci. U. S. A.*, **115**, 6644–6649, doi:10.1073/pnas.1718031115. <http://www.ncbi.nlm.nih.gov/pubmed/29891651> (Accessed April 4, 2019).
- Tighe, M., R. E. Haling, R. J. Flavel, and I. M. Young, 2012: Ecological Succession, Hydrology and Carbon Acquisition of Biological Soil Crusts Measured at the Micro-Scale. *PLoS One*, **7**, e48565, doi:10.1371/journal.pone.0048565. <http://dx.plos.org/10.1371/journal.pone.0048565> (Accessed June 7, 2018).
- Tilman, D., and M. Clark, 2014: Global diets link environmental sustainability and human health. *Nature*, **515**, 518–522, doi:10.1038/nature13959. <http://www.nature.com/articles/nature13959> (Accessed May 30, 2018).

- , K. G. Cassman, P. A. Matson, R. Naylor, and S. Polasky, 2002: Agricultural sustainability and intensive production practices. *Nature*, **418**, 671–677, doi:10.1038/nature01014. <http://www.nature.com/articles/nature01014> (Accessed April 16, 2019).
- , C. Balzer, J. Hill, and B. L. Belfort, 2011: Global food demand and the sustainable intensification of agriculture. *Proc. Natl. Acad. Sci. U. S. A.*, **108**, 20260–20264, doi:10.1073/pnas.1116437108. <http://www.ncbi.nlm.nih.gov/pubmed/22106295> (Accessed May 30, 2018).
- Timmer, C., 2009: Preventing Food Crises Using a Food Policy Approach. *J. Nutr.*, **140**, 224S–228S. <https://academic.oup.com/jn/article-abstract/140/1/224S/4743317> (Accessed November 11, 2018).
- Timmermann, C., and Z. Robaey, 2016: Agrobiodiversity Under Different Property Regimes. *J. Agric. Environ. Ethics*, **29**, 285–303, doi:10.1007/s10806-016-9602-2. <http://link.springer.com/10.1007/s10806-016-9602-2> (Accessed June 1, 2018).
- Tirado, M. C., R. Clarke, L. A. Jaykus, A. McQuatters-Gollop, and J. M. Frank, 2010: Climate change and food safety: A review. *Clim. Chang. Food Sci.*, doi:10.1016/j.foodres.2010.07.003.
- Tirivayi, N., M. Knowles, and B. Davis, 2016: The interaction between social protection and agriculture: A review of evidence. *Glob. Food Sec.*, **10**, 52–62, doi:10.1016/J.GFS.2016.08.004. <https://www.sciencedirect.com/science/article/abs/pii/S2211912416300359> (Accessed April 15, 2019).
- Tom, M. S., P. S. Fischbeck, and C. T. Hendrickson, 2016: Energy use, blue water footprint, and greenhouse gas emissions for current food consumption patterns and dietary recommendations in the US. *Environ. Syst. Decis.*, **36**, 92–103, doi:10.1007/s10669-015-9577-y. <http://link.springer.com/10.1007/s10669-015-9577-y> (Accessed April 15, 2019).
- Torlesse, H., L. Kiess, and M. W. Bloem, 2003a: Association of Household Rice Expenditure with Child Nutritional Status Indicates a Role for Macroeconomic Food Policy in Combating Malnutrition. *J. Nutr.*, **133**, 1320–1325, doi:10.1093/jn/133.5.1320. <https://academic.oup.com/jn/article/133/5/1320/4558593> (Accessed April 15, 2019).
- Torlesse, H., L. Kiess, M. B.-T. J. of Nutrition, and U. 2003, 2003b: Association of household rice expenditure with child nutritional status indicates a role for macroeconomic food policy in combating malnutrition. *J. Nutr.*, **133**, 1320–1325. <https://academic.oup.com/jn/article-abstract/133/5/1320/4558593> (Accessed November 11, 2018).
- Torvanger, A., 2018: Governance of bioenergy with carbon capture and storage (BECCS): accounting, rewarding, and the Paris agreement. *Clim. Policy*, **0**, 1–13, doi:10.1080/14693062.2018.1509044. <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1509044>.
- Trabucco, A., R. J. Zomer, D. A. Bossio, O. van Straaten, and L. V. Verchot, 2008: Climate change mitigation through afforestation/reforestation: A global analysis of hydrologic impacts with four case studies. *Agric. Ecosyst. Environ.*, doi:10.1016/j.agee.2008.01.015.
- Tschakert, P., B. van Oort, A. L. St. Clair, and A. LaMadrid, 2013: Inequality and transformation analyses: a complementary lens for addressing vulnerability to climate change. *Clim. Dev.*, **5**, 340–350, doi:10.1080/17565529.2013.828583. <http://www.tandfonline.com/doi/abs/10.1080/17565529.2013.828583> (Accessed November 10, 2018).
- Tscharntke, T., A. M. Klein, A. Kruess, I. Steffan-Dewenter, and C. Thies, 2005: Landscape perspectives on agricultural intensification and biodiversity – ecosystem service management. *Ecol. Lett.*, **8**, 857–874, doi:10.1111/j.1461-0248.2005.00782.x. <http://doi.wiley.com/10.1111/j.1461-0248.2005.00782.x> (Accessed April 15, 2019).
- , and Coauthors, 2016: When natural habitat fails to enhance biological pest control – Five

- hypotheses. *Biol. Conserv.*, **204**, 449–458, doi:10.1016/J.BIOCON.2016.10.001. <https://www.sciencedirect.com/science/article/abs/pii/S0006320716305249> (Accessed April 15, 2019).
- Tu, J., Z.-G. Xia, K. C. Clarke, and A. Frei, 2007: Impact of Urban Sprawl on Water Quality in Eastern Massachusetts, USA. *Environ. Manage.*, **40**, 183–200, doi:10.1007/s00267-006-0097-x. <http://link.springer.com/10.1007/s00267-006-0097-x> (Accessed April 15, 2019).
- Tullberg, J., D. Antille, C. Bluett, ... J. E.-S. and T., and undefined 2018, Controlled traffic farming effects on soil emissions of nitrous oxide and methane. *Elsevier*.
- Turner, B., 2011: Embodied connections: sustainability, food systems and community gardens. *Local Environ.*, **16**, 509–522, doi:10.1080/13549839.2011.569537. <http://www.tandfonline.com/doi/abs/10.1080/13549839.2011.569537> (Accessed April 15, 2019).
- Uçkun Kiran, E., A. P. Trzcinski, W. J. Ng, and Y. Liu, 2014: Bioconversion of food waste to energy: A review. *Fuel*, **134**, 389–399, doi:10.1016/J.FUEL.2014.05.074. <https://www.sciencedirect.com/science/article/pii/S0016236114005365> (Accessed April 16, 2019).
- UNCCD, 2012: *Zero Net Land Degradation. A sustainable Development Goal for Rio+20. To secure the contribution of our planet's land and soil to sustainable development, including food security and poverty eradication.*
- , 2017: *Global Land Outlook.* Bonn, Germany,.
- UNCTAD, 2011: Water for Food – Innovative Water Management Technologies for Food Security and Poverty Alleviation. 39. http://unctad.org/en/docs/dtlstict2011d2_en.pdf.
- UNEP, 2017: *The Emissions Gap Report 2017.* United Nations Environment Programme (UNEP), Nairobi,.
- UNEP and WMO, 2011: Integrated Assessment of Black Carbon and Tropospheric Ozone: Summary for Decision Makers. United Nations Environment Programme and World Meteorological Organization. http://www.wmo.int/pages/prog/arep/gaw/documents/BlackCarbon_SDM.pdf
- UNWater, 2015: Water for a Sustainable World.
- Upton, B., R. Miner, M. Spinney, and L. S. Heath, 2008: The greenhouse gas and energy impacts of using wood instead of alternatives in residential construction in the United States. *Biomass and Bioenergy*, **32**, 1–10, doi:10.1016/J.BIOMBIOE.2007.07.001. <https://www.sciencedirect.com/science/article/pii/S0961953407001109> (Accessed April 15, 2019).
- Valdez, Z. P., W. C. Hockaday, C. A. Masiello, M. E. Gallagher, and G. Philip Robertson, 2017: Soil Carbon and Nitrogen Responses to Nitrogen Fertilizer and Harvesting Rates in Switchgrass Cropping Systems. *Bioenergy Res.*, doi:10.1007/s12155-016-9810-7.
- Valendik, E. N., S. V. Vershovets, E. K. Kisilyahov, G. A. Ivanova, A. V. Bruchanov, I. V. Kosov, and J. G. Goldammer, 2011: *Tekhnologii kontroliruyemykh vyzhiganiy v lesakh Sibiri: kollektivnaya monografiya [Technologies of controlled burning in forests of Siberia]*. E.S. Petrenko, Ed. Siberian Federal University, Krasnoyarsk, 160 pp.
- Valin, H., and Coauthors, 2014: The future of food demand: understanding differences in global economic models. *Agric. Econ.*, doi:10.1111/agec.12089.
- VandenBygaart, A. J., 2016: The myth that no-till can mitigate global climate change. *Agric. Ecosyst. Environ.*, **216**, 98–99, doi:10.1016/J.AGEE.2015.09.013. <https://www.sciencedirect.com/science/article/pii/S0167880915300797> (Accessed June 5, 2018).

- Vasconcelos, A. C. F., M. Bonatti, S. L. Schlindwein, L. R. D'Agostini, L. R. Homem, and R. Nelson, 2013: Landraces as an adaptation strategy to climate change for smallholders in Santa Catarina, Southern Brazil. *Land use policy*, **34**, 250–254, doi:10.1016/J.LANDUSEPOL.2013.03.017. <https://www.sciencedirect.com/science/article/pii/S026483771300063X> (Accessed June 1, 2018).
- Vatsala, L., J. Prakash, and S. Prabhavathi, 2017: Food security and nutritional status of women selected from a rural area in South India. *J. Food, Nutr. Popul. Heal.*, **1**, 10.
- Vaughan, N. E., and C. Gough, 2016: Expert assessment concludes negative emissions scenarios may not deliver. *Environ. Res. Lett.*, **11**, 95003, doi:10.1088/1748-9326/11/9/095003. <http://stacks.iop.org/1748-9326/11/i=9/a=095003?key=crossref.857f5dec1289a7ab6a51049ce522720c> (Accessed November 11, 2018).
- Vellakkal, S., J. Fledderjohann, S. Basu, S. Agrawal, S. Ebrahim, O. Campbell, P. Doyle, and D. Stuckler, 2015: Food Price Spikes Are Associated with Increased Malnutrition among Children in Andhra Pradesh, India. *J. Nutr.*, **145**, 1942–1949, doi:10.3945/jn.115.211250. <https://academic.oup.com/jn/article/145/8/1942/4585797> (Accessed April 15, 2019).
- Vente, J. de, M. Reed, L. Stringer, S. Valente, and J. Newig, 2016: How does the context and design of participatory decision making processes affect their outcomes? Evidence from sustainable land management in global drylands. *Ecosyst. Serv.*, **21**. <https://www.jstor.org/stable/26270377> (Accessed November 10, 2018).
- Vermeulen, S. J., B. M. Campbell, and J. S. I. Ingram, 2012: Climate Change and Food Systems. *Annu. Rev. Environ. Resour.*, **37**, 195–222, doi:10.1146/annurev-environ-020411-130608. <http://www.annualreviews.org/doi/10.1146/annurev-environ-020411-130608> (Accessed November 8, 2018).
- Vignola, R., C. A. Harvey, P. Bautista-Solis, J. Avelino, B. Rapidel, C. Donatti, and R. Martinez, 2015: Ecosystem-based adaptation for smallholder farmers: Definitions, opportunities and constraints. *Agric. Ecosyst. Environ.*, **211**, 126–132, doi:10.1016/J.AGEE.2015.05.013. <https://www.sciencedirect.com/science/article/pii/S0167880915002157> (Accessed June 1, 2018).
- Vilà, M., and Coauthors, 2011: Ecological impacts of invasive alien plants: a meta-analysis of their effects on species, communities and ecosystems. *Ecol. Lett.*, **14**, 702–708, doi:10.1111/j.1461-0248.2011.01628.x. <http://doi.wiley.com/10.1111/j.1461-0248.2011.01628.x> (Accessed November 11, 2018).
- Vira, B., C. Wildburger, and S. Mansourian, 2015: *Forests and Food: Addressing Hunger and Nutrition Across Sustainable Landscapes*. B. Vira, C. Wildburger, and S. Mansourian, Eds. Open Book Publishers, Cambridge, UK, 280 pp.
- Vlontzos, G., S. Niavis, and B. Manos, 2014: A DEA approach for estimating the agricultural energy and environmental efficiency of EU countries. *Renew. Sustain. Energy Rev.*, **40**, 91–96, doi:10.1016/J.RSER.2014.07.153. <https://www.sciencedirect.com/science/article/pii/S1364032114006054> (Accessed June 1, 2018).
- Vogel, C., and K. O'Brien, 2006: Who can eat information? Examining the effectiveness of seasonal climate forecasts and regional climate-risk management strategies. *Clim. Res.*, doi:10.3354/cr033111.
- Vranken, L., T. Avermaete, D. Petalios, and E. Mathijs, 2014: Curbing global meat consumption: Emerging evidence of a second nutrition transition. *Environ. Sci. Policy*, **39**, 95–106, doi:10.1016/J.ENVSCI.2014.02.009. <https://www.sciencedirect.com/science/article/pii/S1462901114000562> (Accessed April 15, 2019).

- de Vries, W., and Coauthors, 2009: The impact of nitrogen deposition on carbon sequestration by European forests and heathlands. *For. Ecol. Manage.*, **258**, 1814–1823. <https://www.sciencedirect.com/science/article/pii/S0378112709001479> (Accessed November 11, 2018).
- DE VRIES, W., G. J. REINDS, P. GUNDERSEN, and H. STERBA, 2006: The impact of nitrogen deposition on carbon sequestration in European forests and forest soils. *Glob. Chang. Biol.*, **12**, 1151–1173, doi:10.1111/j.1365-2486.2006.01151.x. <http://doi.wiley.com/10.1111/j.1365-2486.2006.01151.x> (Accessed November 10, 2018).
- van Vuuren, D. P., and Coauthors, 2011: RCP2.6: exploring the possibility to keep global mean temperature increase below 2 degrees C. *Clim. Change*, doi:10.1007/s10584-011-0152-3.
- van Vuuren, D. P., and Coauthors, 2015a: Pathways to achieve a set of ambitious global sustainability objectives by 2050: Explorations using the IMAGE integrated assessment model. *Technol. Forecast. Soc. Change*, **98**, 303–323, doi:10.1016/j.techfore.2015.03.005.
- , and Coauthors, 2015b: Pathways to achieve a set of ambitious global sustainability objectives by 2050: Explorations using the IMAGE integrated assessment model. *Technol. Forecast. Soc. Change*, **98**, 303–323, doi:10.1016/j.techfore.2015.03.005. <https://linkinghub.elsevier.com/retrieve/pii/S0040162515000645> (Accessed November 11, 2018).
- van Vuuren, D. P., H. van Soest, K. Riahi, and E. Al., 2016: Carbon budgets and energy transition pathways. *Environ. Res. Lett.*, doi:10.1088/1748-9326/11/7/075002.
- van Vuuren, D. P., and Coauthors, 2017a: Energy, land-use and greenhouse gas emissions trajectories under a green growth paradigm. *Glob. Environ. Chang.*, **42**, 237–250, doi:10.1016/J.GLOENVCHA.2016.05.008. <https://www.sciencedirect.com/science/article/pii/S095937801630067X> (Accessed April 14, 2019).
- , A. F. Hof, M. A. E. van Sluisveld, and K. Riahi, 2017b: Open discussion of negative emissions is urgently needed. *Nat. Energy*, **2**, 902–904, doi:10.1038/s41560-017-0055-2. <http://www.nature.com/articles/s41560-017-0055-2> (Accessed November 11, 2018).
- , and Coauthors, 2018a: Alternative pathways to the 1.5 °C target reduce the need for negative emission technologies. *Nat. Clim. Chang.*, **8**, 391–397, doi:10.1038/s41558-018-0119-8. <http://www.nature.com/articles/s41558-018-0119-8> (Accessed November 11, 2018).
- , and Coauthors, 2018b: Alternative pathways to the 1.5 °C target reduce the need for negative emission technologies. *Nat. Clim. Chang.*, 1–7, doi:10.1038/s41558-018-0119-8.
- Vuuren, D. van, M. Kok, P. Lucas, ... A. P.-... F. and S., and undefined 2015, Pathways to achieve a set of ambitious global sustainability objectives by 2050: explorations using the IMAGE integrated assessment model. *Elsevier*.
- Wada, Y., and M. F. P. Bierkens, 2014a: Sustainability of global water use: Past reconstruction and future projections. *Environ. Res. Lett.*, **9**, doi:10.1088/1748-9326/9/10/104003.
- , and M. F. P. Bierkens, 2014b: Sustainability of global water use: past reconstruction and future projections. *Environ. Res. Lett.*, **9**, 104003, doi:10.1088/1748-9326/9/10/104003. <http://stacks.iop.org/1748-9326/9/i=10/a=104003?key=crossref.5e9e54a6dcd7140e8f70e367028e1217> (Accessed November 10, 2018).
- Waha, K., M. T. van Wijk, S. Fritz, L. See, P. K. Thornton, J. Wichern, and M. Herrero, 2018: Agricultural diversification as an important strategy for achieving food security in Africa. *Glob. Chang. Biol.*, **24**, 3390–3400, doi:10.1111/gcb.14158. <http://doi.wiley.com/10.1111/gcb.14158> (Accessed November 9, 2018).

- Waldron, A., D. Garrity, Y. Malhi, C. Girardin, D. C. Miller, and N. Seddon, 2017: Agroforestry Can Enhance Food Security While Meeting Other Sustainable Development Goals. *Trop. Conserv. Sci.*, **10**, 194008291772066, doi:10.1177/1940082917720667. <http://journals.sagepub.com/doi/10.1177/1940082917720667> (Accessed April 12, 2019).
- Wallace, K. J., 2007: Classification of ecosystem services: Problems and solutions. *Biol. Conserv.*, **139**, 235–246, doi:10.1016/J.BIOCON.2007.07.015. <https://www.sciencedirect.com/science/article/pii/S0006320707002765> (Accessed April 4, 2019).
- Walter, K., A. Don, and H. Flessa, 2015: No general soil carbon sequestration under Central European short rotation coppices. *GCB Bioenergy*, doi:10.1111/gcbb.12177.
- Wang, M., and Coauthors, 2017: On the long-term hydroclimatic sustainability of perennial bioenergy crop expansion over the United States. *J. Clim.*, doi:10.1175/JCLI-D-16-0610.1.
- Wardle, J., K. Parmenter, and J. Waller, 2000a: Nutrition knowledge and food intake. *Appetite*, **34**, 269–275, doi:10.1006/APPE.1999.0311. <https://www.sciencedirect.com/science/article/pii/S0195666399903112> (Accessed June 7, 2018).
- , ———, and ———, 2000b: Nutrition knowledge and food intake. *Appetite*, **34**, 269–275, doi:10.1006/appe.1999.0311. <http://linkinghub.elsevier.com/retrieve/pii/S0195666399903112> (Accessed April 18, 2019).
- Warren, A., 2002: Land degradation is contextual. *L. Degrad. Dev.*, **13**, 449–459, doi:10.1002/ldr.532. <http://doi.wiley.com/10.1002/ldr.532> (Accessed November 11, 2018).
- Watson, J. E. M., N. Dudley, D. B. Segan, and M. Hockings, 2014: The performance and potential of protected areas. *Nature*, **515**, 67–73, doi:10.1038/nature13947.
- Watson, R., and Coauthors, 2011: UK National Ecosystem Assessment: understanding nature’s value to society. Synthesis of key findings.
- Wattnem, T., 2016: Seed laws, certification and standardization: outlawing informal seed systems in the Global South. *J. Peasant Stud.*, **43**, 850–867, doi:10.1080/03066150.2015.1130702. <https://www.tandfonline.com/doi/full/10.1080/03066150.2015.1130702> (Accessed August 10, 2018).
- WBA, 2016: *WBA Global Bioenergy Statistics 2016*.
- Wei, F., S. Wang, L. Zhang, C. Fu, and E. M. Kanga, 2018: Balancing community livelihoods and biodiversity conservation of protected areas in East Africa. *Curr. Opin. Environ. Sustain.*, **33**, 26–33, doi:10.1016/J.COSUST.2018.03.013. <https://www.sciencedirect.com/science/article/pii/S1877343517302804> (Accessed April 4, 2019).
- Weinberger, K., and T. A. Lumpkin, 2007: Diversification into Horticulture and Poverty Reduction: A Research Agenda. *World Dev.*, **35**, 1464–1480, doi:10.1016/J.WORLDDEV.2007.05.002. <https://www.sciencedirect.com/science/article/abs/pii/S0305750X07000769> (Accessed April 15, 2019).
- Weindl, I., H. Lotze-Campen, A. Popp, C. Müller, P. Havlík, M. Herrero, C. Schmitz, and S. Rolinski, 2015: Livestock in a changing climate: production system transitions as an adaptation strategy for agriculture Related content Reducing greenhouse gas emissions in agriculture without compromising food security? Implications of climate mitigation for future agr. *Environ. Res. Lett.*, **10**. <http://iopscience.iop.org/article/10.1088/1748-9326/10/9/094021/pdf>.
- , and Coauthors, 2017: Livestock and human use of land: Productivity trends and dietary choices as drivers of future land and carbon dynamics. *Glob. Planet. Change*, **159**, 1–10, doi:10.1016/j.gloplacha.2017.10.002.

- West, J. J., and Coauthors, 2013: Co-benefits of mitigating global greenhouse gas emissions for future air quality and human health. *Nat. Clim. Chang.*, **3**, 885–889, doi:10.1038/nclimate2009. <http://www.nature.com/articles/nclimate2009> (Accessed June 1, 2018).
- Westerling, A., H. Hidalgo, D. Cayan, and T. Swetnam, 2006: Warming and Earlier Spring Increase Western U.S. Forest Wildfire Activity. *Science*, **313**, 940–943, doi:10.1126/science.262.5135.885. <http://www.ncbi.nlm.nih.gov/pubmed/17757357> (Accessed May 31, 2018).
- Westman, W., 1977: How Much Are Nature’s Services Worth? *Science (80-.)*, **197**, 960–964. https://www.jstor.org/stable/1744285?casa_token=caPhe0BPlaUAAAAA:-75vDW1eE33Ky6HvqUPskVQzHpVzkwfSvpJVe-ItbYTgcJoqvOGw0io5EGWHX5crgLAodCzMR3crN4xJDA6wV71b1PXJL6h746UZbct6JKpZWURK8nV2&seq=1#metadata_info_tab_contents (Accessed April 5, 2019).
- Wheeler, T., and J. von Braun, 2013: Climate Change Impacts on Global Food Security. *Science (80-.)*, doi:10.1126/science.1239402.
- Whitaker, J., and Coauthors, 2018: Consensus, uncertainties and challenges for perennial bioenergy crops and land use. *GCB Bioenergy*, doi:10.1111/gcbb.12488.
- Whitehead, P. J., P. Purdon, J. Russell-Smith, P. M. Cooke, and S. Sutton, 2008: The management of climate change through prescribed Savanna burning: Emerging contributions of indigenous people in Northern Australia. *Public Adm. Dev.*, **28**, 374–385, doi:10.1002/pad.512. <http://doi.wiley.com/10.1002/pad.512> (Accessed May 31, 2018).
- Whitfield, L., 2012: How Countries Become Rich and Reduce Poverty: A Review of Heterodox Explanations of Economic Development. *Dev. Policy Rev.*, **30**, 239–260, doi:10.1111/j.1467-7679.2012.00575.x. <http://doi.wiley.com/10.1111/j.1467-7679.2012.00575.x> (Accessed April 15, 2019).
- Whittinghill, L. J., and D. B. Rowe, 2012: The role of green roof technology in urban agriculture. *Renew. Agric. Food Syst.*, **27**, 314–322.
- Wiebe, K., and Coauthors, 2015: Climate change impacts on agriculture in 2050 under a range of plausible socioeconomic and emissions scenarios. *Environ. Res. Lett.*, **10**, 85010, doi:10.1088/1748-9326/10/8/085010. <http://stacks.iop.org/1748-9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b> (Accessed June 7, 2018).
- Wijitkosum, S., 2016: The impact of land use and spatial changes on desertification risk in degraded areas in Thailand. *Sustain. Environ. Res.*, **26**, 84–92, doi:https://doi.org/10.1016/j.serj.2015.11.004. <http://www.sciencedirect.com/science/article/pii/S246820391630019X>.
- Wild, M., A. Roesch, and C. Ammann, 2012: Global dimming and brightening-Evidence and agricultural implications. *CAB Rev.*, **7**, doi:10.1079/PAVSNR20127003. <https://www.researchgate.net/publication/228534246> (Accessed August 11, 2018).
- Wilhelm, M., C. Blome, V. Bhakoo, and A. Paulraj, 2016: Sustainability in multi-tier supply chains: Understanding the double agency role of the first-tier supplier. *J. Oper. Manag.*, **41**, 42–60. <https://www.sciencedirect.com/science/article/pii/S0272696315300115> (Accessed May 1, 2018).
- Wilhite, D. A., 2005: *Drought and Water Crises*. D. Wilhite, Ed. CRC Press, <https://www.taylorfrancis.com/books/9781420028386> (Accessed April 15, 2019).
- Wilkie, D. S., J. F. Carpenter, and Q. Zhang, 2001: The under-financing of protected areas in the Congo Basin: so many parks and so little willingness-to-pay. *Biodivers. Conserv.*, **10**, 691–709, doi:10.1023/A:1016662027017. <http://link.springer.com/10.1023/A:1016662027017> (Accessed April 4, 2019).

- Wiloso, E. I., R. Heijungs, G. Huppes, and K. Fang, 2016: Effect of biogenic carbon inventory on the life cycle assessment of bioenergy: Challenges to the neutrality assumption. *J. Clean. Prod.*, doi:10.1016/j.jclepro.2016.03.096.
- Wise, M., and Coauthors, 2009a: Implications of limiting CO₂ concentrations for land use and energy. *Science*, **324**, 1183–1186, doi:10.1126/science.1168475. <http://www.ncbi.nlm.nih.gov/pubmed/19478180> (Accessed April 15, 2019).
- Wise, M., and Coauthors, 2009b: Implications of Limiting CO₂ Concentrations for Land Use and Energy. *Science* (80-.), **324**, 1183–1186, doi:10.1126/science.1168475. <http://www.sciencemag.org/cgi/doi/10.1126/science.1168475> (Accessed November 11, 2018).
- Wise, M., E. L. Hodson, B. K. Mignone, L. Clarke, S. Waldhoff, and P. Luckow, 2015: An approach to computing marginal land use change carbon intensities for bioenergy in policy applications. *Energy Econ.*, **50**, 337–347, doi:10.1016/J.ENECO.2015.05.009. <https://www.sciencedirect.com/science/article/pii/S0140988315001619> (Accessed April 5, 2019).
- Wodon, Q., and H. Zaman, 2010: Higher Food Prices in Sub-Saharan Africa: Poverty Impact and Policy Responses. *World Bank Res. Obs.*, **25**, 157–176, doi:10.1093/wbro/lkp018. <https://academic.oup.com/wbro/article-lookup/doi/10.1093/wbro/lkp018> (Accessed June 1, 2018).
- Wolff, S., E. A. Schrammeijer, C. J. E. Schulp, and P. H. Verburg, 2018: Meeting global land restoration and protection targets: What would the world look like in 2050? *Glob. Environ. Chang.*, **52**, 259–272, doi:10.1016/j.gloenvcha.2018.08.002.
- Wollenberg, E., and Coauthors, 2016a: Reducing emissions from agriculture to meet the 2 °C target. *Glob. Chang. Biol.*, **22**, 3859–3864, doi:10.1111/gcb.13340. <http://doi.wiley.com/10.1111/gcb.13340> (Accessed May 30, 2018).
- , and Coauthors, 2016b: Reducing emissions from agriculture to meet the 2 °C target. *Glob. Chang. Biol.*, **22**, 3859–3864, doi:10.1111/gcb.13340. <http://doi.wiley.com/10.1111/gcb.13340> (Accessed April 18, 2019).
- Wong, V. N. L., R. S. B. Greene, R. C. Dalal, and B. W. Murphy, 2010: Soil carbon dynamics in saline and sodic soils: a review. *Soil Use Manag.*, **26**, 2–11, doi:10.1111/j.1475-2743.2009.00251.x. <http://doi.wiley.com/10.1111/j.1475-2743.2009.00251.x> (Accessed November 11, 2018).
- Woolf, D., J. E. Amonette, F. A. Street-Perrott, J. Lehmann, and S. Joseph, 2010: Sustainable biochar to mitigate global climate change. *Nat. Commun.*, **1**, 1–9, doi:10.1038/ncomms1053. <http://www.nature.com/doi/10.1038/ncomms1053> (Accessed May 30, 2018).
- World Bank, FAO, and IFAD, 2009: Module 15: Gender and Forestry. *Gender in Agriculture Sourcebook*, The International Bank for Reconstruction and Development/The World Bank, Washington D.C., 643–674.
- Worrall, F., M. G. Evans, A. Bonn, M. S. Reed, D. Chapman, and J. Holden, 2009: Can carbon offsetting pay for upland ecological restoration? *Sci. Total Environ.*, **408**, 26–36, doi:10.1016/J.SCITOTENV.2009.09.022. <https://www.sciencedirect.com/science/article/pii/S0048969709008663> (Accessed August 13, 2018).
- Wright, C. K., and M. C. Wimberly, 2013: Recent land use change in the Western Corn Belt threatens grasslands and wetlands. *Proc. Natl. Acad. Sci.*, **110**, 4134–4139, doi:10.1073/pnas.1215404110. <http://www.pnas.org/cgi/doi/10.1073/pnas.1215404110> (Accessed November 11, 2018).
- Wu, J., 1999: Crop Insurance, Acreage Decisions, and Nonpoint-Source Pollution. *Am. J. Agric. Econ.*, **81**, 305, doi:10.2307/1244583. <https://academic.oup.com/ajae/article-lookup/doi/10.2307/1244583> (Accessed April 15, 2019).

- Wu, W., and Coauthors, 2019: Global advanced bioenergy potential under environmental protection policies and societal transformation measures. *GCB Bioenergy*, **Accepted**.
- Xie, H., and C. Ringler, 2017: Agricultural nutrient loadings to the freshwater environment: the role of climate change and socioeconomic change. *Environ. Res. Lett.*, **12**, 104008, doi:10.1088/1748-9326/aa8148. <http://stacks.iop.org/1748-9326/12/i=10/a=104008?key=crossref.6fb7bd42a49a14001a406f9311d8c26d> (Accessed April 13, 2019).
- Xu, D., A. Song, D. Li, X. Ding, and Z. Wang, 2018: Assessing the relative role of climate change and human activities in desertification of North China from 1981 to 2010. *Front. Earth Sci.*, doi:10.1007/s11707-018-0706-z.
- Xu, Y., and Ramanathan, V., 2017: Well below 2 °C: Mitigation strategies for avoiding dangerous to catastrophic climate changes, PNAS, 114(39), 10315-10323, doi: 10.1073/pnas.1618481114. <https://doi.org/10.1073/pnas.1618481114>
- Xu, Y., Zaelke, D., Velders, G. J. M., and Ramanathan, V., 2013: The role of HFCs in mitigating 21st century climate change, Atmos. Chem. Phys., 13, 6083–6089, doi:10.5194/acp-13-6083-2013. <https://doi.org/10.5194/acp-13-6083-2013>
- Yamagata, Y., N. Hanasaki, A. Ito, T. Kinoshita, D. Murakami, and Q. Zhou, 2018: Estimating water–food–ecosystem trade-offs for the global negative emission scenario (IPCC-RCP2.6). *Sustain. Sci.*, **0**, 0, doi:10.1007/s11625-017-0522-5. <http://link.springer.com/10.1007/s11625-017-0522-5>.
- Yang, H., and A. J. . Zehnder, 2002: Water Scarcity and Food Import: A Case Study for Southern Mediterranean Countries. *World Dev.*, **30**, 1413–1430, doi:10.1016/S0305-750X(02)00047-5. <https://www.sciencedirect.com/science/article/abs/pii/S0305750X02000475> (Accessed April 15, 2019).
- Yang, L., L. Chen, W. Wei, Y. Yu, H. Z.-J. of Hydrology, and undefined 2014, Comparison of deep soil moisture in two re-vegetation watersheds in semi-arid regions. *Elsevier*,
- Yirdaw, E., M. Tigabu, and A. Monge, 2017: Rehabilitation of degraded dryland ecosystems – review. *Silva Fenn.*, doi:10.14214/sf.1673.
- Yong, D., and K. Peh, 2016: South-east Asia’s forest fires: blazing the policy trail. *Oryx*, **50**, 207–212. <https://www.cambridge.org/core/journals/oryx/article/southeast-asias-forest-fires-blazing-the-policy-trail/CA89A255B7D7DA310F2F02268178D33A> (Accessed November 11, 2018).
- Young, C. E., and P. C. Westcott, 2000: *How Decoupled Is U.S. Agricultural Support for Major Crops?* 762-767 pp. <http://ajae.oxfordjournals.org/> (Accessed April 15, 2019).
- Zaehle, S., and D. Dalmonech, 2011: Carbon–nitrogen interactions on land at global scales: current understanding in modelling climate biosphere feedbacks. *Curr. Opin. Environ. Sustain.*, **3**, 311–320. <https://www.sciencedirect.com/science/article/pii/S1877343511000868> (Accessed November 11, 2018).
- Zaman, A. U., and S. Lehmann, 2011: Urban growth and waste management optimization towards “zero waste city.” *City, Cult. Soc.*, **2**, 177–187, doi:10.1016/J.CCS.2011.11.007. <https://www.sciencedirect.com/science/article/pii/S1877916611000786> (Accessed April 15, 2019).
- , and ——, 2013: The zero waste index: a performance measurement tool for waste management systems in a “zero waste city.” *J. Clean. Prod.*, **50**, 123–132, doi:10.1016/J.JCLEPRO.2012.11.041. <https://www.sciencedirect.com/science/article/pii/S095965261200635X> (Accessed April 15, 2019).

- Zeza, A., G. Carletto, B. Davis, K. Stamoulis, and P. Winters, 2009: Rural income generating activities: whatever happened to the institutional vacuum? Evidence from Ghana, Guatemala, Nicaragua and Vietnam. *World Dev.*, **37**, 1297–1306. <https://www.sciencedirect.com/science/article/pii/S0305750X08003306> (Accessed November 11, 2018).
- Zhang, K., Y. Zhang, H. Tian, X. Cheng, H. Dang, and Q. Zhang, 2013: Sustainability of social–ecological systems under conservation projects: Lessons from a biodiversity hotspot in western China. *Biol. Conserv.*, **158**, 205–213, doi:<https://doi.org/10.1016/j.biocon.2012.08.021>.
- Zhang, X., R. Obringer, C. Wei, N. Chen, and D. Niyogi, 2017: Droughts in India from 1981 to 2013 and Implications to Wheat Production. *Sci. Rep.*, **7**, 44552, doi:10.1038/srep44552. <http://www.nature.com/articles/srep44552> (Accessed June 2, 2018).
- Zhou, G., X. Zhou, Y. He, J. Shao, Z. Hu, R. Liu, H. Zhou, and S. Hosseinibai, 2017: Grazing intensity significantly affects belowground carbon and nitrogen cycling in grassland ecosystems: a meta-analysis. *Glob. Chang. Biol.*, **23**, 1167–1179, doi:10.1111/gcb.13431. <http://doi.wiley.com/10.1111/gcb.13431> (Accessed November 12, 2018).
- Ziervogel, G., and P. J. Ericksen, 2010: Adapting to climate change to sustain food security. *Wiley Interdiscip. Rev. Clim. Chang.*, **1**, 525–540, doi:10.1002/wcc.56. <http://doi.wiley.com/10.1002/wcc.56> (Accessed November 11, 2018).
- Zomer, R., H. Neufeldt, J. Xu, A. Ahrends, D. Bossio, A. Trabucco, M. Van Noordwijk, and M. Wang, 2016: Global Tree Cover and Biomass Carbon on Agricultural Land: The contribution of agroforestry to global and national carbon budgets. *Sci. Rep.*, **6**, 29987. <https://www.nature.com/articles/srep29987> (Accessed November 10, 2018).

1 **Chapter 7: Risk management and decision making in relation to** 2 **sustainable development**

3
4 **Coordinating Lead Authors:** Margot Hurlbert (Canada), Jagdish Krishnaswamy (India)

5 **Lead Authors:** Edouard Davin (France/Switzerland), Francis X. Johnson (Sweden), Carlos Fernando
6 Mena (Ecuador), John Morton (United Kingdom), Soojeong Myeong (The Republic of Korea), David
7 Viner (United Kingdom), Koko Warner (The United States of America), Anita Wreford (New
8 Zealand), Sumaya Zakieldean (Sudan), Zinta Zommers (Latvia)

9 **Contributing Authors:** Rob Bailis (The United States of America), Brigitte Baptiste (Colombia),
10 Kerry Bowman (Canada), Edward Byers (Austria/Brazil), Katherine Calvin (The United States of
11 America), Rocio Diaz-Chavez (Mexico), Jason Evans (Australia), Amber Fletcher (Canada), James
12 Ford (United Kingdom), Sean Patrick Grant (The United States of America), Darshini Mahadevia
13 (India), Yousef Manialawy (Canada), Pamela McElwee (The United States of America), Minal Pathak
14 (India), Julian Quan (United Kingdom), Balaji Rajagopalan (The United States of America), Alan
15 Renwick (New Zealand), Jorge E. Rodríguez-Morales (Peru), Charlotte Streck (Germany), Wim
16 Thiery (Belgium), Alan Warner (Barbados)

17 **Review Editors:** Regina Rodrigues (Brazil), B.L. Turner II (The United States of America)

18 **Chapter Scientist:** Thobekile Zikhali (Zimbabwe)

19 **Date of Draft:** 07/08/2019

| | | |
|----|---|-----------|
| 1 | Table of Contents | |
| 2 | Chapter 7: Risk management and decision making in relation to sustainable development | 1 |
| 3 | Executive summary..... | 4 |
| 4 | 7.1. Introduction and Relation to Other Chapters | 9 |
| 5 | 7.1.1. Findings of Previous IPCC Assessments and Reports..... | 9 |
| 6 | 7.1.2. Treatment of Key Terms in the Chapter | 10 |
| 7 | 7.1.3. Roadmap to the chapter..... | 11 |
| 8 | 7.2. Climate-related risks for land-based human systems and ecosystems..... | 11 |
| 9 | 7.2.1. Assessing Risk | 12 |
| 10 | 7.2.2. Risks to land systems arising from climate change | 12 |
| 11 | 7.2.3. Risks arising from responses to climate change | 19 |
| 12 | 7.2.4. Risks arising from Hazard, Exposure, and Vulnerability..... | 22 |
| 13 | 7.3. Consequences of climate – land change for human well-being and sustainable development | |
| 14 | 27 | |
| 15 | 7.3.1. What is at stake for food security?..... | 27 |
| 16 | 7.3.2. Risks to where and how people live: Livelihood systems and migration | 27 |
| 17 | 7.3.3. Risks to humans from disrupted ecosystems and species | 28 |
| 18 | 7.3.4. Risks to Communities and Infrastructure..... | 29 |
| 19 | Cross-chapter Box 10: Economic dimensions of climate change and land | 30 |
| 20 | 7.4. Policy Instruments for Land and Climate | 33 |
| 21 | 7.4.1. Multi-level Policy Instruments..... | 34 |
| 22 | 7.4.2. Policies for Food Security and Social Protection..... | 37 |
| 23 | 7.4.3. Policies Responding to Climate Related Extremes | 40 |
| 24 | 7.4.4. Policies Responding to GHG fluxes | 43 |
| 25 | 7.4.5. Policies Responding to Desertification and Degradation – Land Degradation Neutrality | |
| 26 | (LDN) 48 | |
| 27 | 7.4.6. Policies Responding to Land Degradation..... | 50 |
| 28 | 7.4.7. Economic and financial instruments for adaptation, mitigation, and land..... | 57 |
| 29 | 7.4.8. Enabling effective policy instruments – Policy Portfolio Coherence | 60 |
| 30 | 7.4.9. Barriers to Implementing Policy Responses | 62 |
| 31 | Cross-chapter Box 11: Gender in inclusive approaches to climate change, land, and sustainable | |
| 32 | development..... | 66 |
| 33 | 7.5. Decision-making for Climate Change and Land..... | 69 |
| 34 | 7.5.1. Formal and Informal decision-making..... | 69 |
| 35 | 7.5.2. Decision Making, Timing, Risk, and Uncertainty | 71 |
| 36 | 7.5.3. Best practices of decision making toward sustainable land management | 74 |
| 37 | 7.5.4. Adaptive management | 75 |

| | | |
|----|---|-----|
| 1 | 7.5.5. Performance indicators | 77 |
| 2 | 7.5.6. Maximising Synergies and Minimising Trade-offs | 78 |
| 3 | 7.6. Governance: Governing the land-climate interface | 93 |
| 4 | 7.6.1. Institutions Building Adaptive and Mitigative Capacity..... | 93 |
| 5 | 7.6.2. Integration - Levels, Modes, and Scale of Governance for Sustainable Development. | 95 |
| 6 | Cross-Chapter Box 12: Traditional biomass use: land, climate and development implications | 99 |
| 7 | 7.6.3. Adaptive Climate Governance Responding to Uncertainty | 101 |
| 8 | 7.6.4. Participation | 106 |
| 9 | Cross-Chapter Box 13: Indigenous and Local Knowledge in the IPCC Special Reports | 107 |
| 10 | 7.6.5. Land Tenure | 111 |
| 11 | 7.6.6. Institutional dimensions of adaptive governance | 117 |
| 12 | 7.6.7. Inclusive Governance for Sustainable Development | 118 |
| 13 | 7.7. Key uncertainties and knowledge gaps | 119 |
| 14 | Frequently Asked Questions | 120 |
| 15 | References..... | 122 |
| 16 | Supplementary Material..... | 233 |
| 17 | | |

1 **Executive summary**

2 **Increases in global mean surface temperature are projected to result in continued permafrost**
3 **degradation and coastal degradation (*high confidence*), increased wildfire, decreased crop yields**
4 **in low latitudes, decreased food stability, decreased water availability, vegetation loss (*medium***
5 ***confidence*), decreased access to food and increased soil erosion (*low confidence*). There is *high***
6 ***agreement and high evidence* that increases in global mean temperature will result in continued**
7 **increase in global vegetation loss, coastal degradation, as well as decreased crop yields in low**
8 **latitudes, decreased food stability, decreased access to food and nutrition, and *medium***
9 ***confidence* in continued permafrost degradation and water scarcity in drylands. Impacts are**
10 **already observed across all components (*high confidence*). Some processes may experience**
11 **irreversible impacts at lower levels of warming than others. There are high risks from permafrost**
12 **degradation, and wildfire, coastal degradation, stability of food systems at 1.5°C while high risks from**
13 **soil erosion, vegetation loss and changes in nutrition only occur at higher temperature thresholds due**
14 **to increased possibility for adaptation (*medium confidence*). {7.2.2.1, 7.2.2.2, 7.2.2.3; 7.2.2.4; 7.2.2.5;**
15 **7.2.2.6; 7.2.2.7; Figure 7.1}**

16
17 **These changes result in compound risks to food systems, human and ecosystem health,**
18 **livelihoods, the viability of infrastructure, and the value of land (*high confidence*). The**
19 **experience and dynamics of risk change over time as a result of both human and natural processes**
20 **(*high confidence*). There is *high confidence* that climate and land changes pose increased risks at**
21 **certain periods of life (i.e. to the very young and ageing populations) as well as sustained risk to those**
22 **living in poverty. Response options may also increase risks. For example, domestic efforts to insulate**
23 **populations from food price spikes associated with climatic stressors in the mid-2000s inadequately**
24 **prevented food insecurity and poverty, and worsened poverty globally. {7.2.1, 7.2.2, 7.3, Table 7.1}**

25
26 **There is significant regional heterogeneity in risks: tropical regions, including Sub-Saharan**
27 **Africa, Southeast Asia and Central and South America are particularly vulnerable to decreases**
28 **in crop yield (*high confidence*). Yield of crops in higher latitudes may initially benefit from warming**
29 **as well as from higher CO₂ concentrations. But temperate zones, including the Mediterranean, North**
30 **Africa, the Gobi desert, Korea and western United States are susceptible to disruptions from increased**
31 **drought frequency and intensity, dust storms and fires (*high confidence*). {7.2.2}**

32
33 **Risks related to land degradation, desertification and food security increase with temperature**
34 **and can reverse development gains in some socio-economic development pathways (*high***
35 ***confidence*) . SSP1 reduces the vulnerability and exposure of human and natural systems and**
36 **thus limits risks resulting from desertification, land degradation and food insecurity compared**
37 **to SSP3 (*high confidence*). SSP1 is characterized by low population growth, reduced inequalities,**
38 **land use regulation, low meat consumption, increased trade and few barriers to adaptation or**
39 **mitigation. SSP3 has the opposite characteristics. Under SSP1, only a small fraction of the dryland**
40 **population (around 3% at 3°C for the year 2050) will be exposed and vulnerable to water stress.**
41 **However under SSP3, around 20% of dryland populations (for the year 2050) will be exposed and**
42 **vulnerable to water stress by 1.5°C and 24% by 3°C. Similarly under SSP1, at 1.5°C, 2 million people**
43 **are expected to be exposed and vulnerable to crop yield change. Over 20 million are exposed and**
44 **vulnerable to crop yield change in SSP3, increasing to 854 million people at 3°C (*low confidence*).**
45 **Livelihoods deteriorate as a result of these impacts, livelihood migration is accelerated, and strife and**
46 **conflict is worsened (*medium confidence*). {Cross-Chapter Box 9 in Chapter 6, 7.2.2, 7.3.2, Table 7.1,**
47 **Figure 7.2}**

48

1 **Land-based adaptation and mitigation responses pose risks associated with the effectiveness and**
2 **potential adverse side-effects of measures chosen (*high confidence*).** Adverse side-effects on food
3 security, ecosystem services and water security increase with the scale of bioenergy and bioenergy
4 with carbon capture and storage (BECCS) deployment. In a SSP1 future, bioenergy and BECCS
5 deployment up to 6 Mkm² is compatible with sustainability constraints, whereas risks are already high
6 in a SSP3 future for this scale of deployment. {7.2.3}

7
8 **There is *high confidence* that policies addressing vicious cycles of poverty, land degradation and**
9 **greenhouse gas emissions implemented in a holistic manner can achieve climate resilient**
10 **sustainable development. Choice and implementation of policy instruments determine future**
11 **climate and land pathways (*medium confidence*).** Sustainable development pathways (described in
12 SSP1) supported by effective regulation of land use to reduce environmental trade-offs, reduced
13 reliance on traditional biomass, low growth in consumption and limited meat diets, moderate
14 international trade with connected regional markets, and effective GHG mitigation instruments) can
15 result in lower food prices, fewer people affected by floods and other climatic disruptions, and
16 increases in forested land (*high agreement, limited evidence*) (SSP1). A policy pathway with limited
17 regulation of land use, low technology development, resource intensive consumption, constrained
18 trade, and ineffective GHG mitigation instruments can result in food price increases, and significant
19 loss of forest (*high agreement, limited evidence*) (SSP3). {3.7.5, 7.2.2, 7.3.4, 7.5.5, 7.5.6, Table 7.1,
20 Cross-Chapter Box 12: Traditional Biomass, in this chapter}

21
22 **Delaying deep mitigation in other sectors and shifting the burden to the land sector, increases**
23 **the risk associated with adverse effects on food security and ecosystem services(*high confidence*).**
24 The consequences are an increased pressure on land with higher risk of mitigation failure and of
25 temperature overshoot and a transfer of the burden of mitigation and unabated climate change to
26 future generations. Prioritising early decarbonisation with minimal reliance on carbon dioxide
27 removal (CDR) decreases the risk of mitigation failure (*high confidence*). {2.5, 6.2, 6.4, 7.2.1, 7.2.2,,
28 7.2.3, 7.5.6, 7.5.7, Cross-Chapter Box 9 in Chapter 6, 7.5.6}

29
30 **Trade-offs can occur between using land for climate mitigation or sustainable development goal**
31 **(SDG) 7 (affordable clean energy) with biodiversity, food, ground-water and riverine ecosystem**
32 **services (*medium confidence*).** There is *medium confidence* that trade-offs currently do not figure
33 into climate policies and decision making. Small hydro power installations (especially in clusters) can
34 impact downstream river ecological connectivity for fish (*high agreement, medium evidence*). Large
35 scale solar farms and wind turbine installations can impact endangered species and disrupt habitat
36 connectivity (*medium agreement, medium evidence*). Conversion of rivers for transportation can
37 disrupt fisheries and endangered species (through dredging and traffic) (*medium agreement, low*
38 *evidence*). {7.5.6}

39
40 **The full mitigation potential assessed in this report will only be realised if agricultural emissions**
41 **are included in mainstream climate policy (*high agreement, high evidence*) .** Carbon markets are
42 theoretically more cost-effective than taxation but challenging to implement in the land-sector (*high*
43 *confidence*) Carbon pricing (through carbon markets or carbon taxes) has the potential to be an
44 effective mechanism to reduce GHG emissions, although it remains relatively untested in agriculture
45 and food systems. Equity considerations can be balanced by a mix of both market and non-market
46 mechanisms (*medium evidence, medium agreement*). Emissions leakage could be reduced by multi-
47 lateral action (*high agreement, medium evidence*). {7.4.6, 7.5.5, 7.5.6, Cross Chapter Box 9 in
48 Chapter 6}

1 **A suite of coherent climate and land policies advances the goal of the Paris Agreement and the**
2 **land-related SDG targets on poverty, hunger, health, sustainable cities and communities,**
3 **responsible consumption and production, and life on land. There is *high confidence* that acting**
4 **early will avert or minimise risks, reduce losses and generate returns on investment** . The
5 economic costs of action on sustainable land management, mitigation, and adaptation are less than the
6 consequences of inaction for humans and ecosystems (*medium confidence*). Policy portfolios that
7 make ecological restoration more attractive, people more resilient - expanding financial inclusion,
8 flexible carbon credits, disaster risk and health insurance, social protection and adaptive safety nets,
9 contingent finance and reserve funds, and universal access to early warning systems – could save
10 USD 100 billion a year, if implemented globally. {7.3.1, 7.4.7, 7.4.8, 7.5.6, Cross-chapter box 10:
11 Economic Dimensions, in this chapter}

12
13 **Coordination of policy instruments across scales, levels, and sectors advances co-benefits,**
14 **manages land and climate risks, advances food security, and addresses equity concerns (*medium***
15 ***confidence*).** Flood resilience policies are mutually reinforcing and include flood zone mapping,
16 financial incentives to move, and building restrictions, and insurance. Sustainability certification,
17 technology transfer, land use standards and secure land tenure schemes, integrated with early action
18 and preparedness, advance response options. Sustainable land management improves with investment
19 in agricultural research, environmental farm practices, agri-environmental payments, financial support
20 for sustainable agricultural water infrastructure (including dugouts), agriculture emission trading, and
21 elimination of agricultural subsidies (*medium confidence*). Drought resilience policies (including
22 drought preparedness planning, early warning and monitoring, improving water use efficiency),
23 synergistically improve agricultural producer livelihoods and foster sustainable land management.
24 {3.7.5, Cross-Chapter Box 5 in Chapter 3, 7.4.3, 7.4.6, 7.5.6, 7.4.8, , 7.5.6, 7.6.3}

25
26 **Technology transfer in land use sectors offers new opportunities for adaptation, mitigation,**
27 **international cooperation, R&D collaboration, and local engagement (*medium confidence*).**
28 International cooperation to modernise the traditional biomass sector will free up both land and labour
29 for more productive uses. Technology transfer can assist the measurement and accounting of emission
30 reductions by developing countries. {7.4.4, 7.4.6}

31
32 **Measuring progress towards goals is important in decision-making and adaptive governance to**
33 **create common understanding and advance policy effectiveness (*high agreement, medium***
34 ***evidence*).** Measurable indicators, selected with the participation of people and supporting data
35 collection, are useful for climate policy development and decision-making. Indicators include the
36 SDGs, nationally determined contributions (NDCs), land degradation neutrality (LDN) core
37 indicators, carbon stock measurement, measurement and monitoring for REDD+, metrics for
38 measuring biodiversity and ecosystem services, and governance capacity. {7.5.5, 7.5.7, 7.6.4, 7.6.6}

39
40 **The complex spatial, cultural and temporal dynamics of risk and uncertainty in relation to land**
41 **and climate interactions and food security, require a flexible, adaptive, iterative approach to**
42 **assessing risks, revising decisions and policy instruments (*high confidence*).** Adaptive, iterative
43 decision making moves beyond standard economic appraisal techniques to new methods such as
44 dynamic adaptation pathways with risks identified by trigger points through indicators. Scenarios can
45 provide valuable information at all planning stages in relation to land, climate and food; adaptive
46 management addresses uncertainty in scenario planning with pathway choices made and reassessed to
47 respond to new information and data as it becomes available. {3.7.5, 7.4.4, 7.5.2, 7.5.3, 7.5.4, 7.5.7,
48 7.6.1, 7.6.3}

1 **Indigenous and local knowledge (ILK) can play a key role in understanding climate processes**
2 **and impacts, adaptation to climate change, sustainable land management across different**
3 **ecosystems, and enhancement of food security** (*high confidence*). ILK is context-specific,
4 collective, informally transmitted, and multi-functional, and can encompass factual information about
5 the environment and guidance on management of resources and related rights and social behaviour.
6 ILK can be used in decision-making at various scales and levels, and exchange of experiences with
7 adaptation and mitigation that include ILK is both a requirement and an entry strategy for
8 participatory climate communication and action. Opportunities exist for integration of ILK with
9 scientific knowledge. {7.4.1, 7.4.5, 7.4.6, 7.6.4, Cross-Chapter Box 13: in this chapter}

10
11 **Participation of people in land and climate decision making and policy formation allows for**
12 **transparent effective solutions and the implementation of response options that advance**
13 **synergies, reduce trade-offs in sustainable land management** (*high confidence*), **and overcomes**
14 **barriers to adaptation and mitigation** (*high confidence*). Improvements to sustainable land
15 management are achieved by: (1) engaging people in citizen science by mediating and facilitating
16 landscape conservation planning, policy choice, and early warning systems (*medium confidence*); (2)
17 involving people in identifying problems (including species decline, habitat loss, land use change in
18 agriculture, food production and forestry), selection of indicators, collection of climate data, land
19 modelling, agricultural innovation opportunities. When social learning is combined with collective
20 action, transformative change can occur addressing tenure issues and changing land use practices
21 (*medium confidence*). Meaningful participation overcomes barriers by opening up policy and science
22 surrounding climate and land decisions to inclusive discussion that promotes alternatives. {3.7.5,
23 7.4.1, 7.4.9; 7.5.1, 7.5.4, 7.5.5, 7.5.7, 7.6.4, 7.6.6}

24
25 **Empowering women can bolster synergies among household food security and sustainable land**
26 **management** (*high confidence*). This can be achieved with policy instruments that account for
27 gender differences. The overwhelming presence of women in many land based activities including
28 agriculture provides opportunities to mainstream gender policies, overcome gender barriers, enhance
29 gender equality, and increase sustainable land management and food security (*high confidence*).
30 Policies that address barriers include gender qualifying criteria and gender appropriate delivery,
31 including access to financing, information, technology, government transfers, training, and extension
32 may be built into existing women's programs, structures (civil society groups) including collective
33 micro enterprise (*medium confidence*) . {Cross-Chapter Box 11 in this chapter}

34
35 **The significant social and political changes required for sustainable land use, reductions in**
36 **demand and land-based mitigation efforts associated with climate stabilisation require a wide**
37 **range of governance mechanisms**. The expansion and diversification of land use and biomass
38 systems and markets requires hybrid governance: public-private partnerships, transnational,
39 polycentric, and state governance to insure opportunities are maximised, trade-offs are managed
40 equitably and negative impacts are minimised (*medium confidence*). {7.4.6, 7.6.2, 7.6.3, Cross-
41 Chapter Box 7 in Chapter 6}

42
43 **Land tenure systems have implications for both adaptation and mitigation, which need to be**
44 **understood within specific socio-economic and legal contexts, and may themselves be impacted**
45 **by climate change and climate action** (*limited evidence, high agreement*). Land policy (in a
46 diversity of forms beyond focus on freehold title) can provide routes to land security and facilitate or
47 constrain climate action, across cropping, rangeland, forest, fresh-water ecosystems and other
48 systems. Large-scale land acquisitions are an important context for the relations between tenure
49 security and climate change, but their scale, nature and implications are imperfectly understood. There
50 is *medium confidence* that land titling and recognition programs, particularly those that authorize and

1 respect indigenous and communal tenure, can lead to improved management of forests, including for
2 carbon storage. Strong public coordination (government and public administration) can integrate land
3 policy with national policies on adaptation and reduce sensitivities to climate change. {7.6.2; 7.6.3;
4 7.6.4, 7.6.5}

5
6 **Significant gaps in knowledge exist when it comes to understanding the effectiveness of policy**
7 **instruments and institutions related to land use management, forestry, agriculture and**
8 **bioenergy. Interdisciplinary research is needed on the impacts of policies and measures in land**
9 **sectors.** Knowledge gaps are due in part to the highly contextual and local nature of land and climate
10 measures and the long time periods needed to evaluate land use change in its socio-economic frame,
11 as compared to technological investments in energy or industry that are somewhat more comparable.
12 Significant investment is needed in monitoring, evaluation and assessment of policy impacts across
13 different sectors and levels. {7.7}

14
15

7.1. Introduction and Relation to Other Chapters

Land is integral to human habitation and livelihoods, providing food and resources, and also serves as a source of identity and cultural meaning. However, the combined impacts of climate change, desertification, land degradation and food insecurity pose obstacles to resilient development and the achievement of the Sustainable Development Goals (SDGs). This chapter reviews and assesses literature on risk and uncertainty surrounding land and climate change, policy instruments and decision-making addressing those risks and uncertainty, and governance practices that advance response options with co-benefits identified in Chapter 6, lessen the socio-economic impacts of climate change and reduce trade-offs, and advance sustainable land management.

7.1.1. Findings of Previous IPCC Assessments and Reports

This chapter builds on earlier assessments contained in several chapters of the IPCC Fifth Assessment Report (the contributions of both Working Groups II and III), the IPCC Special Report on Managing the Risks of Extreme Events and Disasters to Advance Climate Change Adaptation (SREX)(IPCC 2012), and the IPCC Special Report on Global Warming of 1.5°C (SR15). (IPCC 2018a) The findings most relevant to decision-making on and governance of responses to land-climate challenges are set out in Box 7.1.

Box 7.1 Relevant Findings of Recent IPCC Reports

Climate change and sustainable development pathways

“Climate change poses a moderate threat to current sustainable development and a severe threat to future sustainable development” (Denton et al. 2014; Fleurbaey et al. 2014).

Significant transformations may be required for climate-resilient pathways (Denton et al. 2014; Jones et al. 2014).

The design of climate policy is influenced by: (1) differing ways that individuals and organisations perceive risks and uncertainties; (2) the consideration of a diverse array of risks and uncertainties as well as human and social responses which may be difficult to measure, are of low probability but which would have a significant impact if they occurred (Kunreuther et al. 2014; Fleurbaey et al. 2014; Kolstad et al. 2014).

Building climate resilient pathways requires iterative, continually evolving and complementary processes at all levels of government (Denton et al. 2014; Kunreuther et al. 2014; Kolstad et al. 2014; Somanthan et al. 2014; Lavell et al. 2012).

Important aspects of climate resilient policies include local level institutions, decentralisation, participatory governance, iterative learning, integration of local knowledge, and reduction of inequality (Dasgupta et al. 2014; Lavell et al. 2012; Cutter et al. 2012b; O’ Brien et al. 2012; Roy et al. 2018).

Climate action and sustainable development are linked: adaptation has co-benefits for sustainable development while “sustainable development supports, and often enables, the fundamental societal and systems transitions and transformations that help limit global warming” (IPCC 2018b). Redistributive policies that shield the poor and vulnerable can resolve trade-offs between mitigation objectives and the hunger, poverty and energy access SDGs.

Land and rural livelihoods

Policies and institutions relating to land, including land tenure, can contribute to the vulnerability of rural people, and constrain adaptation. Climate policies, such as encouraging cultivation of biofuels,

1 or payments under REDD+, will have significant secondary impacts, both positive and negative, in
2 some rural areas (Dasgupta et al. 2014).

3 “Sustainable land management is an effective disaster risk reduction tool”(Cutter et al. 2012a).

4 **Risk and risk management**

5 A variety of emergent risks not previously assessed or recognised, can be identified by taking into
6 account: a) the “interactions of climate change impacts on one sector with changes in exposure and
7 vulnerability, as well as adaptation and mitigation actions”, and; b) “indirect, trans-boundary, and
8 long-distance impacts of climate change” including price spikes, migration, conflict and the
9 unforeseen impacts of mitigation measures (Oppenheimer et al. 2014)

10 “Under any plausible scenario for mitigation and adaptation, some degree of risk from residual
11 damages is unavoidable” (Oppenheimer et al. 2014).

12 **Decision-making**

13 “Risk management provides a useful framework for most climate change decision-making. Iterative
14 risk management is most suitable in situations characterised by large uncertainties, long time frames,
15 the potential for learning over time, and the influence of both climate as well as other socioeconomic
16 and biophysical changes” (Jones et al. 2014).

17 “Decision support is situated at the intersection of data provision, expert knowledge, and human
18 decision making at a range of scales from the individual to the organisation and institution” (Jones et
19 al. 2014).

20 “Scenarios are a key tool for addressing uncertainty”, either through problem exploration or solution
21 exploration (Jones et al. 2014).

22 **Governance**

23 There is no single approach to adaptation planning and both top-down and bottom-up approaches are
24 widely recognised. “Institutional dimensions in adaptation governance play a key role in promoting
25 the transition from planning to implementation of adaptation” (Mimura et al. 2014). Adaptation is also
26 essential at all scales, including adaptation by local governments, businesses, communities, and
27 individuals (Denton et al. 2014).

28 “Strengthened multi-level governance, institutional capacity, policy instruments, technological
29 innovation and transfer and mobilisation of finance, and changes in human behaviour and lifestyles
30 are enabling conditions that enhance the feasibility of mitigation and adaptation options for 1.5°C –
31 consistent systems transitions” (IPCC 2018b).

32 Governance is key for vulnerability and exposure represented by institutionalised rule systems and
33 habitualised behaviour and norms that govern society and guide actors and , “it is essential to improve
34 knowledge on how to promote adaptive governance within the framework of risk assessment and risk
35 management” (Cardona 2012).

37 **7.1.2. Treatment of Key Terms in the Chapter**

38 While the term risk continues to be subject to a growing number of definitions in different disciplines
39 and sectors, this chapter takes as a starting point the definition used in the IPCC Special Report on
40 Global Warming of 1.5°C (SR15) (IPCC 2018a), which reflects definitions used by both Working
41 Group II and Working Group III in the Fifth Assessment Report (AR5): “The potential for adverse
42 consequences where something of value is at stake and where the occurrence and degree of an
43 outcome is uncertain” (Allwood et al. 2014; Oppenheimer et al. 2014). The SR15 definition further
44 specifies: “In the context of the assessment of climate impacts, the term risk is often used to refer to
45 the potential for adverse consequences of a climate-related hazard, or of adaptation or mitigation
46 responses to such a hazard, on lives, livelihoods, health and wellbeing, ecosystems and species,

1 economic, social and cultural assets, services (including ecosystem services), and infrastructure”. In
2 SR15, as in the IPCC SREX and AR5 WGII, risk is conceptualised as resulting from the interaction of
3 vulnerability (of the affected system), its exposure over time (to a hazard), as well as the (climate-
4 related) impact and the likelihood of its occurrence (AR5 2014; IPCC 2018a, 2012). In the context of
5 SRCCL, risk must also be seen as including risks to the implementation of responses to land-climate
6 challenges from economic, political and governance factors. Climate and land risks must be seen in
7 relation to human values and objectives (Denton et al. 2014). Risk is closely associated with concepts
8 of vulnerability and resilience, which are themselves subject to differing definitions across different
9 knowledge communities.

10 Risks examined in this chapter arise from more than one of the major land-climate-society challenges
11 (desertification, land degradation, and food insecurity), or partly stem from mitigation or adaptation
12 actions, or cascade across different sectors or geographical locations. They could thus be seen as
13 examples of *emergent risks* (Oppenheimer et al. 2014, p. 1052): “aris[ing] from the interaction of
14 phenomena in a complex system”. Stranded assets in the coal sector due to proliferation of renewable
15 energy and government response could be examples of emergent risks (Saluja, N and Singh 2018;
16 Marcacci 2018). Additionally, the absence of an explicit goal for conserving fresh-water ecosystems
17 and ecosystem services in SDGs (in contrast to a goal (Life Under Water) that is exclusively for
18 marine biodiversity) is related to its trade-offs with energy and irrigation goals thus posing a
19 substantive risk (Nilsson et al. 2016b; Vörösmarty et al. 2010).

20 *Governance* is not previously well defined in IPCC reports, but is used here to include all of the
21 processes, structures, rules and traditions that govern, which may be undertaken by actors including
22 governments, markets, organisations, or families (Bevir 2011), with particular reference to the
23 multitude of actors operating in respect of land and climate interactions. Such definitions of
24 governance allow for it to be decoupled from the more familiar concept of government and studied in
25 the context of complex human-environment relations and environmental and resource regimes (Young
26 2017a). Governance involves the interactions among formal and informal institutions through which
27 people articulate their interests, exercise their legal rights, meet their legal obligations, and mediate
28 their differences (UNDP 1997).

29 **7.1.3. Roadmap to the chapter**

30 This chapter firstly discusses risks and their drivers, at various scales, in relation to land-climate
31 challenges, including risks associated with responses to climate change (Section 7.2). The
32 consequences of the principal risks in economic and human terms, and associated concepts such as
33 tipping points and windows of opportunity for response are then described (Section 7.3). Policy
34 responses at different scales to different land-climate risks, and barriers to implementation, are
35 described in Section 7.4, followed by assessment of approaches to decision-making on land-climate
36 challenges (Section 7.5), and questions of the governance of the land-climate interface (Section 7.6).
37 Key uncertainties and knowledge gaps are identified (Section 7.7).

38 **7.2. Climate-related risks for land-based human systems and** 39 **ecosystems**

40 This section examines risks that climate change pose to selected land-based human systems and
41 ecosystems, and then further explores how social and economic choices, as well as responses to
42 climate change, will exacerbate or lessen risks. Risk is the potential for adverse consequences for
43 human or ecological systems, recognising the diversity of values and objectives associated with such
44 systems. The interacting processes of climate change, land change, and unprecedented social and
45 technological change, pose significant risk to climate resilient sustainable development. The pace,
46 intensity, and scale of these sizeable risks affect the central issues in sustainable development: access
47 to ecosystem services and resources essential to sustain people in given locations, how and where

1 people live and work, and the means to safeguard human wellbeing against disruptions (Warner et al.
2 2019). In the context of climate change, adverse consequences can arise from the potential *impacts of*
3 climate change as well as human *responses to* climate change. Relevant adverse consequences include
4 those on lives, livelihoods, health and wellbeing, economic, social and cultural assets and
5 investments, infrastructure, services (including ecosystem services), ecosystems and species (see
6 Glossary). Risks result from dynamic interactions between climate-related hazards with the exposure
7 and vulnerability of the affected human or ecological system to the hazards. Hazards, exposure and
8 vulnerability may change over time and space as a result of socio-economic changes and human
9 decision-making (*risk management*). Numerous uncertainties exist in the scientific understanding of
10 risk (See Chapter 1.2.2).

11 **7.2.1. Assessing Risk**

12 This chapter applies and further improves methods used in previous IPCC reports including AR5 and
13 the Special Report on Global Warming of 1.5° (SR15) to assess risks. Evidence is drawn from
14 published studies, which include observations of impacts from human-induced climate change and
15 model projections for future climate change. Such projections are based on IAMs, ESMs, regional
16 climate models and global or regional impact models examining the impact of climate change on
17 various indicators (see Cross-Chapter Box 1: Scenarios, in Chapter 1). Results of laboratory and field
18 experiments that examine impacts of specific changes were also included in the review. Risks under
19 differed future socio-economic conditions were assessed using recent publications based on Shared
20 Socio-economic pathways (SSPs). SSPs provide storylines about future socio-economic development
21 and can be combined with RCPs (Riahi et al. 2017)(see Cross-Chapter Box 9: Illustrative climate and
22 land pathways, in Chapter 6). Risk arising from land-based mitigation and adaptation choices is
23 assessed using studies examining the adverse side-effects of such responses (7.2.3).

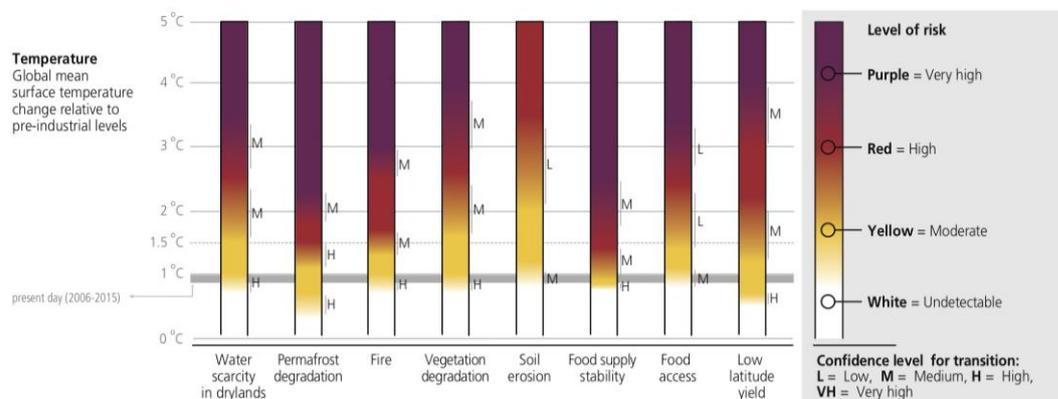
24
25 Burning embers figures introduced in the IPCC Third Assessment Report through to the Fifth
26 Assessment Report, and the SR15, were developed for this report to illustrate risks at different
27 temperature thresholds. Key components involved in desertification, land degradation and food
28 security were identified based on discussions with authors in Chapter 3 –5. The final list of burning
29 embers in Figure 7.1 is not intended to be fully comprehensive, but represents processes for which
30 sufficient literature exists to make expert judgements. Literature used in the burning embers
31 assessment is summarised in table(s) in supplementary material. Following an approach articulated in
32 O’Neill et al. (2017), expert judgements were made to assess thresholds of risk (O’Neill et al. 2017a).
33 To further strengthen replicability of the method, a predefined protocol based on a modified Delphi
34 process was followed (Mukherjee et al. 2015). This included two separate anonymous rating rounds,
35 feedback in between rounds and a group discussion to achieve consensus.

36
37 Burning embers provide ranges of a given variable (typically global mean near-surface air
38 temperature) for which risks transitions from one risk category to the next. Four categories are
39 considered: undetectable, moderate, high and very high. Moderate risk indicates that impacts are
40 detectable and attributable to climate-related factors. High risk indicates widespread impacts on larger
41 number or proportion of population/ area but with the potential to adapt or recover. Very high risk
42 indicates severe and possibly irreversible impacts with limited ability of societies and ecosystems to
43 adapt to them. Transitions between risk categories were assigned confidence levels based on the
44 amount, and quality, of academic literature supporting judgements: L= Low, M = Medium, and H =
45 High. Further details of the procedure is provided in supplementary material.

46 47 **7.2.2. Risks to land systems arising from climate change**

48 At current levels of global mean surface temperature (GMST) increase, impacts are already detectable
49 across numerous land-related systems (*high confidence*) (see chapters 2, 3, 4, 6). There is *high*
50 *confidence* that unabated future climate change will result in continued changes to processes involved
51 in desertification, land degradation and food security, including: water scarcity in drylands, soil
52 erosion, coastal degradation, vegetation loss, fire, permafrost thaw as well as access, stability,

1 utilisation and physical availability of food (Figure 7.1). These changes will increase risks to food
 2 systems, the health of humans and ecosystems, livelihoods, the value of land, infrastructure and
 3 communities (7.3). Details of the risks, and their transitions, are described in the following
 4 subsections.
 5



6
 7
 8 **Figure 7.1: Risks to selected land system elements as a function of global mean surface temperature**
 9 **increase since pre-industrial times. Impacts on human and ecological systems include: 1) economic loss**
 10 **and declines in livelihoods and ecosystem services from water scarcity in drylands, 2) damage to natural**
 11 **and built environment from permafrost thaw related ground instability, 3) damage to infrastructure,**
 12 **altered land cover, accelerated erosion and increased air pollution from fires, 4) vegetation loss and shifts**
 13 **in vegetation structure, 5) economic loss and declines in livelihoods and ecosystem services from reduced**
 14 **land productivity due to soil erosion, 6) increased disruption of food supply (stability), 7) increased**
 15 **disruption of food access and 8) changes to crop yield and food availability in low-latitude regions. Risks**
 16 **are global (3,4,5,6,7) and specific to certain regions (1,2,8). Selected components are illustrative and not**
 17 **intended to be fully comprehensive of factors influencing food security, land degradation and**
 18 **desertification. The supporting literature is provided in Supplementary Material.**
 19
 20

21 7.2.2.1. Crop yield in low latitudes

22 There is *high confidence* that climate change has resulted in decreases in yield (of wheat, rice, maize,
 23 soy) and reduced food availability in low-latitude regions (IPCC, 2018, 5.2.2). Countries in low-
 24 latitude regions are particularly vulnerable because the livelihoods of high proportions of the
 25 population are dependent on agricultural production. Even moderate temperature increases (1°C to 2
 26 °C) have negative yield impacts for major cereals, because the climate of many tropical agricultural
 27 regions is already quite close to the high temperature thresholds for suitable production of these cereal
 28 (Rosenzweig et al. 2014). Thus, by 1.5°C GMT, or between approximately 1.6°C and approximately
 29 2.6°C of local warming, risks to yields may already transition to high in West Africa, Southeast Asia
 30 and Central and South America (Faye et al. 2018) (*medium confidence*). For further information see
 31 5.3.2.1. By contrast, higher latitudes may initially benefit from warming as well as well higher CO₂
 32 concentrations (IPCC 2018a). Wheat yield losses are expected to be lower for the United States (−5.5
 33 ± 4.4% per degree Celsius) and France (−6.0 ± 4.2% per degree Celsius) compared to India (−9.1 ±
 34 5.4% per degree Celsius) (Zhao et al. 2017). Very high risks to low latitude yields may occur between
 35 3°C and 4°C (*medium confidence*). At these temperatures, catastrophic reductions in crop yields may
 36 occur, of up to 60% in low latitudes (Rosenzweig et al. 2014)(5.2.2, 5.2.3). Some studies report
 37 significant population displacement from the tropics related to systemic livelihood disruption in
 38 agriculture systems (Tittonell 2014; Montaña et al. 2016; Huber-Sannwald et al. 2012; Wise et al.
 39 2016; Tanner et al. 2015; Mohapatra 2013). However, at higher temperatures of warming, all regions
 40 of the world face risks of declining yields as a result of extreme weather events and reduced heat
 41 tolerance of maize, rice, wheat and soy (Zhao et al. 2017; IPCC 2018a).
 42

7.2.2.2. Stability of and access to food supplies

Stability of food supply is expected to decrease as the magnitude and frequency of extreme events increase, disrupting food chains in all areas of the world (Wheeler and Von Braun 2013; Coates 2013; Puma et al. 2015; Deryng et al. 2014; Harvey et al. 2014b; Iizumi et al. 2013; Seaman et al. 2014)(*medium evidence, high agreement*)(5.3.2, 5.3.3, 5.6.2, 5.7.1). While international trade in food is assumed to be a key response for alleviating hunger, historical data and economic models suggest that international trade does not adequately redistribute food globally to offset yield declines or other food shortages when weather extremes reduce crop yields (Schmitz et al. 2012; Chatzopoulos et al. 2019; Marchand et al. 2016; Gilbert 2010; Wellesley et al. 2017) (*medium confidence*). When droughts, heat waves, floods or other extremes destroy crops, evidence has shown key producing countries have constrained exports contributing to price spikes and social tension in importing countries which reduces access to food (von Uexkull et al. 2016; Gleick 2014; Maystadt and Ecker 2014; Kelley et al. 2015; Church et al. 2017; Götz et al. 2013; Puma et al. 2015; Willenbockel 2012; Headey 2011; Distefano et al. 2018; Brooks 2014)(*medium evidence, medium agreement*). There is little understanding of how food system shocks cascade through a modern interconnected economy. Reliance on global markets may reduce some risks, but the on-going globalisation of food trade networks exposes the world food system to new impacts that have not been seen in the past (5.1.2, 5.2.1, 5.5.2.5, 5.6.5, 5.7.1). The global food system is vulnerable to systemic disruptions and increasingly interconnected inter-country food dependencies and changes in frequency and severity of extreme weather events may complicate future responses(Puma et al. 2015; Jones and Hiller 2017).

Impacts of climate change are already detectable on food supply and access as price and trade reactions have occurred in response to heat waves, droughts and other extreme events (Noble et al. 2014; O'Neill et al. 2017b)(*high evidence, high agreement*). The impact of climate change on food stability is underexplored (Schleussner et al. 2016; James et al. 2017). However, some literature assesses that by about 2035, daily maximum temperatures will exceed the 90th percentile of historical (1961–1990) temperatures on 25–30% of days (O'Neill et al. 2017b)(ref 35, Figs 11–17) with negative shocks to food stability and world food prices. O'Neill et al. (2017b) remark that in the future, return periods for precipitation events globally (land only) will reduce from one-in-20-year (historical) to about once-in-14-year or less by 2046–2065 in many areas of the world. Domestic efforts to insulate populations from food price spikes associated with climatic stressors in the mid-2000s have been shown to inadequately shield from poverty, and worsen poverty globally (Diffenbaugh et al. 2012; Meyfroidt et al. 2013; Hertel et al. 2010). The transition to high risk is estimated to occur around 1.4°C, possibly by 2035, due to changes in temperature and heavy precipitation events (*medium confidence*) (O'Neill et al. 2017b; Fritsche et al. 2017a; Harvey et al. 2014b). Very high risk may occur by 2.4°C (*medium confidence*) and 4°C of warming is considered catastrophic (IPCC 2018c; Noble et al. 2014) for food stability and access because a combination of extreme events, compounding political and social factors, and shocks to crop yields can heavily constrain options to ensure food security in import-reliant countries.

7.2.2.3. Soil Erosion

Soil erosion increases risks of economic loss and declines in livelihoods due to reduced land productivity. In the EU, on-site costs of soil erosion by wind has been reported at an average of 55 USD per hectare annually, but up to USD 450 per hectare for sugar beet and oilseed rape (Middleton et al. 2017)). Farmers in the Dapo watershed in Ethiopia lose about USD 220 per hectare of maize due to loss of nitrogen through soil erosion (Erkossa et al. 2015). Soil erosion not only increases crop loss but has been shown to have negative household feeding, with older farmers most vulnerable to losses from erosion (Ighodaro et al. 2016). Erosion also results in increased risks to human health, through air pollution from aerosols (Middleton et al. 2017), and brings risks of reduced ecosystem services including supporting services related to soil formation.

1
2 At current levels of warming, changes in erosion are already detected in many regions. Attribution to
3 climate change is challenging as there are other powerful drivers of erosion (e.g., land use), limited
4 global-scale studies (Li and Fang 2016a; Vanmaercke et al. 2016a) and the absence of formal
5 detection and attribution studies (4.2.3). However, studies have found an increase in short-duration
6 and intensity precipitation, due to anthropogenic climate change, which is a causative factor for soil
7 erosion (Lenderink and van Meijgaard 2008; Li and Fang 2016b). High risks of erosion may occur
8 between 2° and 3.5° (*low confidence*) as continued increases in intense precipitation is projected at
9 these temperature thresholds (Fischer and Knutti 2015) in many regions. Warming also reduces soil
10 organic matter, diminishing resistance against erosion. There is *low confidence* concerning the
11 temperature threshold at which risks become very high due to large regional differences and limited
12 global-scale studies (Li and Fang 2016b; Vanmaercke et al. 2016b) (4.4).

13 14 **7.2.2.4. Dryland water scarcity**

15 Water scarcity in drylands contributes to changes in desertification and hazards such as dust storms,
16 increasing risks of economic loss, declines in livelihoods of communities and negative health effects
17 (*high confidence*) (3.1.3). Further information specific to costs and impacts of water scarcity and
18 droughts is detailed in Cross-Chapter Box 5: Case study on policy response to drought, in Chapter 3.

19
20 The IPCC AR5 report and the SR15 concluded that there is *low confidence* in the direction of drought
21 trends since 1950 at the global scale. While these reports did not assess water scarcity with a specific
22 focus on drylands, they indicated that there is *high confidence* in observed drought increases in some
23 regions of the world, including in the Mediterranean and West Africa (IPCC AR5) and that there is
24 *medium confidence* that anthropogenic climate change has contributed to increased drying in the
25 Mediterranean region (including southern Europe, northern Africa and the Near East) and that this
26 tendency will continue to increase under higher levels of global warming (IPCC 2018d). Some parts
27 of the drylands have experienced decreasing precipitation over recent decades (IPCC AR5; Chapter 3,
28 3.2), consistent with the fact that climate change is implicated in desertification trends in some regions
29 (3.2.2). Dust storms, linked to changes in precipitation and vegetation, appear to be occurring with
30 greater frequency in some deserts and their margins (Goudie 2014) (3.3.1). There is therefore *high*
31 *confidence* that the transition from undetectable to moderate risk associated with water scarcity in
32 drylands occurred in recent decades in the range 0.7°C to 1°C (Fig. 7.1).

33
34 Between 1.5°C and 2.5°C, the risk level is expected to increase from moderate to high (*medium*
35 *confidence*). Globally, at 2°C an additional 8% of the world population (of population in 2000) will be
36 exposed to new or aggravated water scarcity (IPCC 2018d). However, at 2°C, the annual warming
37 over drylands will reach 3.2°C–4.0°C, implying about 44% more warming over drylands than humid
38 lands (Huang et al. 2017), thus potentially aggravating water scarcity issues through increased
39 evaporative demand. (Byers et al. 2018a) estimate that 3–22% of the drylands population (range
40 depending on socio-economic conditions) will be exposed and vulnerable to water stress. The
41 Mediterranean, North Africa and the Levant will be particularly vulnerable to water shortages and
42 expansion of desert terrain and vegetation is predicted to occur in the Mediterranean biome, an
43 unparalleled change in the last 10,000 years (*medium confidence*) (IPCC 2018d). At 2.5°C–3.5°C
44 risks are expected to become very high with migration from some drylands resulting as the only
45 adaptation option (*medium confidence*). Scarcity of water for irrigation is expected to increase, in
46 particular in Mediterranean regions, with limited possibilities for adaptation (Haddeland et al. 2014).

47 48 **7.2.2.5. Vegetation degradation**

49 There are clear links between climate change and vegetation cover changes, tree mortality, forest
50 diseases, insect outbreaks, forest fires, forest productivity and net ecosystem biome production (Allen
51 et al. 2010; Bentz et al. 2010; Anderegg et al. 2013; Hember et al. 2017; Song et al. 2018; Sturrock et
52 al. 2011). Forest dieback, often a result of drought and temperature changes, not only produces risks
53 to forest ecosystems but also to people with livelihoods dependent on forests. A 50 year study of
54 temperate forest, dominated by beech (*Fagus sylvatica* L.), documented a 33% decline in basal area

1 and 70% decline in juvenile tree species, possibly as a result of interacting pressures of drought,
2 overgrazing and pathogens (Martin et al. 2015). There is *high confidence* that such dieback impacts
3 ecosystem properties and services including soil microbial community structure (Gazol et al. 2018).
4 Forest managers and users have reported negative emotional impacts from forest dieback such as
5 pessimism about losses, hopelessness, and fear (Oakes et al. 2016). Practices and policies such as
6 forest classification systems, projection of growth, yield and models for timber supply are already
7 being affected by climate change (Sturrock et al. 2011).

8
9 While risks to ecosystems and livelihoods from vegetation degradation are already detectable at
10 current levels of GMT increase, risks are expected to reach high levels between 1.6°C and 2.6°C
11 (*medium confidence*). Significant uncertainty exists due to countervailing factors: CO₂ fertilisation
12 encourages forest expansion but increased drought, insect outbreaks, and fires result in dieback
13 (Bonan 2008; Lindner et al. 2010). The combined effects of temperature and precipitation change,
14 with CO₂ fertilisation, make future risks to forests very location specific. It is challenging therefore to
15 make global estimates. However, even locally specific studies make clear that very high risks occur
16 between 2.6°C and 4°C (*medium confidence*). Australian tropical rainforests experience significant
17 loss of biodiversity with 3.5°C increase. There are no areas with greater than 30 species and all
18 endemics disappear from low and mid-elevation regions (Williams et al. 2003). Mountain ecosystems
19 are particularly vulnerable (Loarie et al. 2009).

20 21 **7.2.2.6. Fire damage**

22 Increasing fires result in heightened risks to infrastructure, accelerated erosion, altered hydrology,
23 increased air pollution, and negative mental health impacts. Fire not only destroys property but
24 induces changes in underlying site conditions (ground cover, soil water repellency, aggregate stability
25 and surface roughness) which amplifies runoff and erosion, increasing future risks to property and
26 human lives during extreme rainfall events (Pierson and Williams 2016). Dust and ash from fires can
27 impact air quality in a wide area. For example, a dust plume from a fire in Idaho, USA, in September
28 2010 was visible in MODIS satellite imagery and extended at least 100 km downwind of the source
29 area (Wagenbrenner et al. 2013). Individuals can suffer from property damage or direct injury,
30 psychological trauma, depression, post traumatic stress disorder and have reported negative impacts to
31 well being from loss of connection to landscape (Paveglio et al. 2016; Sharples et al. 2016a). Costs of
32 large wildfires in the United States can exceed USD 20 million a day (Pierson et al. 2011) and has
33 been estimated at USD8.5 billion per year in Australia (Sharples et al. 2016b). Globally, human
34 exposure to fire will increase due to projected population growth in fire-prone regions (Knorr et al.
35 2016a).

36
37 It is not clear how quickly, or even if, systems can recover from fires. Longevity of effects may differ
38 depending on cover recruitment rate and soil conditions, recovering in one to two seasons or over ten
39 growing seasons (Pierson et al. 2011). In Russia, one third of forest area affected by fires turned into
40 unproductive areas where natural reforestation is not possible within 2–3 life cycles of major forest
41 forming species (i.e., 300–600 years) (Shvidenko et al. 2012).

42
43 Risks under current warming levels are already moderate as anthropogenic climate change has caused
44 significant increases in fire area (*high confidence*) due to availability of detection and attribution
45 studies) (Cross-Chapter Box 3: Fire and climate change, in Chapter 2). This has been detected and
46 attributed regionally, notably in Western US (Abatzoglou and Williams 2016; Westerling et al. 2006;
47 Dennison et al. 2014), Indonesia (Fernandes et al. 2017) and other regions (Jolly et al. 2015).
48 Regional increases have been observed despite a global-average declining trend induced by human
49 fire suppression strategies especially in savannas (Yang et al. 2014a; Andela et al. 2017).

50
51 High risks of fire may occur between 1.3°C and 1.7°C (*medium confidence*). Studies note heightened
52 risks as “fire weather” and land prone to fire increase above 1.5°C (Abatzoglou et al. 2019a), with
53 *medium confidence* in this transition, due to complex interplay between (i) global warming (ii) CO₂-
54 fertilisation, and (iii) human/economic factors affecting fire risk. Canada, the USA and Mediterranean

1 may be particularly vulnerable as the combination of increased fuel due to CO₂ fertilisation, and
2 weather conditions conducive to fire increase risks to people and property. Some studies show
3 substantial effects at 3°C (Knorr et al. 2016b; Abatzoglou et al. 2019b), indicating a transition to very
4 high risks (*medium confidence*). At high warming levels, climate change may become the primary
5 driver of fire risk in the extratropics (Knorr et al. 2016b; Abatzoglou et al. 2019b; Yang et al. 2014b).
6 Pyroconvection activity may increase, in areas such as southeast Australia (Dowdy and Pepler 2018),
7 posing major challenges to adaptation.
8

9 **7.2.2.7. Permafrost**

10 There is a risk of damage to natural and built environment from permafrost thaw related ground
11 instability. Residential, transportation, and industrial infrastructure in the pan-Arctic permafrost area
12 are particularly at risk (Hjort et al., 2018). High risks already exist at low temperatures (*high*
13 *confidence*). Approximately, 21–37% of Arctic permafrost is projected to thaw under 1.5°C of
14 warming (Hoegh-Guldberg et al., 2018). This increases to very high risk around 2°C (between 1.8 and
15 2.3°C) of temperature increase since pre-industrial times (*medium confidence*) with 35–47% of the
16 Arctic permafrost thawing (Hoegh-Guldberg et al., 2018). If climate stabilised at 2°C, still
17 approximately 40% of permafrost area would be lost (Chadburn et al., 2017), leading to nearly four
18 million people and 70% of current infrastructure in the pan-Arctic permafrost area exposed to
19 permafrost thaw and high hazard (Hjort et al., 2018). Indeed between 2°C and 3°C a collapse of
20 permafrost may occur with a drastic biome shift from tundra to boreal forest (Driyfhout et al. 2015;
21 SR15). There is mixed evidence of a tipping point in permafrost collapse, leading to enhanced
22 greenhouse gas emission and particularly methane, between 2°C and 3°C (Hoegh-Guldberg et al.,
23 2018).
24

25 **7.2.2.8. Risks of desertification, land degradation and food insecurity under** 26 **different Future Development Pathways**

27 Socio-economic developments and policy choices that govern land-climate interactions are an
28 important driver of risk along with climate change (*very high confidence*). Risks under two different
29 Shared Socio-economic Pathways (SSPs) were assessed using emerging literature. SSP1 is
30 characterised by low population growth, reduced inequalities, land-use regulation, low meat
31 consumption, and moderate trade (Riahi et al. 2017; Popp et al. 2017a). SSP3 is characterised by high
32 population growth, higher inequalities, limited land-use regulation, resource-intensive consumption
33 including meat-intensive diets, and constrained trade (for further details see Chapter 1 and Cross-
34 Chapter Box 9: Illustrative climate and land pathways in Chapter 6). These two SSPs, among the set
35 of five SSPs, were selected because they illustrate contrasting futures, ranging from low (SSP1) to
36 high (SSP3) challenges to mitigation and adaptation. Figure 7.2 shows that for a given global mean
37 temperature change, risks are different under SSP1 compared to SSP3. In SSP1, global temperature
38 change does not increase above 3°C even in the baseline case (i.e., with no additional mitigation
39 measures) because in this pathway the combination of low population and autonomous improvements,
40 for example, in terms of carbon intensity and/or energy intensity, effectively act as mitigation
41 measures (Riahi et al., 2017). Thus Figure 7.2 does not indicate risks beyond this point in either SSP1
42 and SSP3. Literature based on such socio-economic and climate models is still emerging and there is a
43 need for greater research on impacts of different pathways. There are few SSP studies exploring
44 aspects of desertification and land degradation, but a greater number of SSP studies on food security
45 (see supplementary material). SSP1 reduces the vulnerability and exposure of human and natural
46 systems and thus limits risks resulting from desertification, land degradation and food insecurity
47 compared to SSP3 (*high confidence*).
48
49
50
51

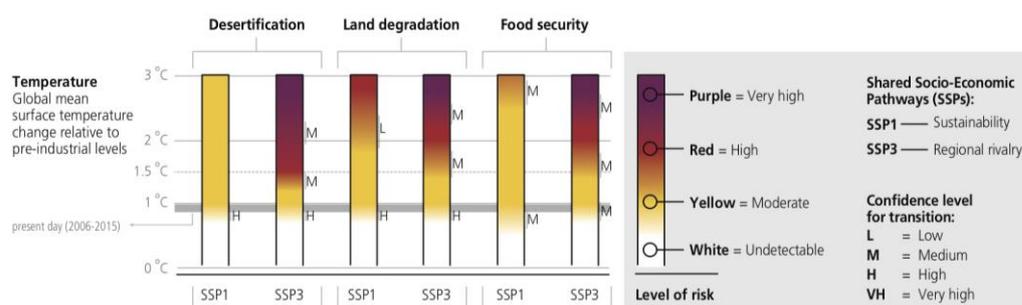


Figure 7.2: Risks associated with desertification, land degradation and food security as a function of climate change and level of socio-economic development. Increasing risks associated to desertification include a growing fraction of population exposed and vulnerable to water scarcity and changes in irrigation supply and demand. Risks related to land degradation include increased vegetation loss, population exposed to fire and floods, costs of floods, extent of deforestation, and ecosystem services including the ability of land to sequester carbon. Risks to food security include population at risk of hunger, food price increases, increases in disability adjusted life years. The risks are assessed for two contrasted socio-economic futures (SSP1 and SSP3) under unmitigated climate change {3.5; 4.2.1.2; 5.2.2; 5.2.3; 5.2.4; 5.2.5; 6.1.4; 7.2}. **The supporting literature is provided in Chapter 7 Supplementary Material.**

Changes to the water cycle due to global warming is an essential driver of desertification and of the risks to livelihood, food production and vegetation in dryland regions. Changes in water scarcity due to climate change have already been detected in some dryland regions (section 7.2.2.4) and therefore the transition to moderate risk occurred in recent decades (*high confidence*). (IPCC 2018d) noted that in the case of risks to water resources, socio-economic drivers are expected to have a greater influence than the changes in climate (*medium confidence*). Indeed, in SSP1 there is only moderate risk even at 3°C of warming, due to the lower exposure and vulnerability of human population (Hanasaki et al. 2013a; Arnell and Lloyd-Hughes 2014; Byers et al. 2018b). Considering drylands only, (Byers et al. 2018b) estimate, using a time sampling approach for climate change and the 2050 population, that at 1.5°C, 2°C and 3°C, the dryland population exposed and vulnerable to water stress in SSP1 will be 2%, 3% and 3% respectively, thus indicating relatively stable moderate risks. In SSP3, the transition from moderate to high risk occurs in the range 1.2°C to 1.5 °C (*medium confidence*) and the transition from high to very high risk is in the range 1.5°C to 2.8 °C (*medium confidence*). (Hanasaki et al. 2013b) found a consistent increase in water stress at higher warming levels due in large part related to growth in population and demand for energy and agricultural commodities and to a lesser extent due to hydrological changes induced by global warming. In SSP3, (Byers et al. 2018b) estimate that at 1.5°C, 2°C and 3°C, the population exposed and vulnerable to water stress in drylands will steadily increase from 20% to 22% and 24%, respectively, thus indicating overall much higher risks compared to SSP1 for the same global warming levels.

SSP studies relevant to land degradation assess risks such as: number of people exposed to fire, the costs of floods and coastal flooding, and loss of ecosystem services including the ability of land to sequester carbon. The risks related to permafrost melting (section 7.2.2.7) are not considered here due to the lack of SSP studies addressing this topic. Climate change impacts on various components of land degradation have already been detected (sections 7.2.2.3; 7.2.2.5; 7.2.2.6) and therefore the transition from undetectable to moderate risk is in the range 0.7 °C –1°C (*high confidence*). Less than 100 million people are exposed to habitat degradation at 1.5°C under SSP1 in non-dryland regions, increasing to 257 million at 2°C (Byers et al. 2018). This suggests a gradual transition to high risk in the range 1.8°C to 2.8°C, but a *low confidence* is attributed due to the very limited evidence to constrain this transition.

1 By contrast in SSP3, there are already 107 million people exposed to habitat degradation at 1.5°C,
2 increasing to 1156 million people at 3°C (Byers et al. 2018b). Furthermore, (Knorr et al. 2016b)
3 estimate that 646 million people will be exposed to fire at 2°C warming, the main risk driver being the
4 high population growth in SSP3 rather than increased burned area due the climate change. Exposure
5 to extreme rainfall, a causative factor for soil erosion and flooding, also differs under SSPs. Under
6 SSP1 up to 14% of the land and population experience five day extreme precipitation events. Similar
7 levels of exposure occur at lower temperatures in SSP3 (Zhang et al. 2018b). Population exposed to
8 coastal flooding is lowest under SSP1 and higher under SSP3 with a limited effect of enhanced
9 protection in SSP3 already after 2°C warming (Hinkel et al. 2014). The transition from high to very
10 high risk will occur at 2.2°C –2.8°C in SSP3 (*medium confidence*), whereas this level of risk is not
11 expected to be reached in SSP1.
12

13 The greatest number of SSP studies explore climate change impacts relevant to food security,
14 including population at risk of hunger, food price increases, increases in disability adjusted life years
15 (Hasegawa et al. 2018a; Wiebe et al. 2015a; van Meijl et al. 2018a; Byers et al. 2018b). Changes in
16 crop yields and food supply stability have already been attributed to climate change (sections 7.2.2.1;
17 7.2.2.2) and the transition from undetectable to moderate risk is placed at 0.5°C – 1°C (*medium*
18 *confidence*). At 1.5°C, about 2 million people are exposed and vulnerable to crop yield change in
19 SSP1 (Hasegawa et al. 2018b; Byers et al. 2018b), implying moderate risk. A transition from
20 moderate to high risk is expected above 2.5°C (*medium confidence*) with population at risk of hunger
21 of the order of 100 million (Byers et al. 2018b). Under SSP3, high risks already exist at 1.5°C
22 (*medium confidence*), with 20 million people exposed and vulnerable to crop yield change. By 2°C,
23 178 million are vulnerable and 854 million people are vulnerable at 3°C (Byers et al. 2018b). This is
24 supported by the higher food prices increase of up to 20% in 2050 in a RCP6.0 scenario (i.e., slightly
25 below 2°C) in SSP3 compared to up to 5% in SSP1 (van Meijl et al. 2018). Furthermore in SSP3,
26 restricted trade increase this price effect (Wiebe et al. 2015). In SSP3, the transition from high to very
27 high risk is in the range 2°C –2.7°C (*medium confidence*) while this transition is never reached in
28 SSP1. This overall confirms that socio-economic development, by affecting exposure and
29 vulnerability, has an even larger effect than climate change for future trends in the population at risk
30 of hunger O'Neill et al. (2017) (p32). Changes can also threaten development gains (*medium*
31 *confidence*). Disability adjusted life years due to childhood underweight decline in both SSP1 and
32 SSP3 by 2030 (by 36.4 million disability adjusted life years in SSP1 and 16.2 million in SSP3).
33 However by 2050, disability adjusted life years increase by 43.7 million in SSP3 (Ishida et al. 2014).
34

35 **7.2.3. Risks arising from responses to climate change**

36 **7.2.3.1. Risk associated with land-based adaptation**

37 Land-based adaptation relates to a particular category of adaptation measures relying on land
38 management (Sanz et al. 2017). While most land-based adaptation options provide co-benefits for
39 climate mitigation and other land challenges (Chapter 6, 6.4.1), in some contexts adaptation measures
40 can have adverse side-effects, thus implying a risk to socio-ecological systems.
41

42 One example of risk is the possible decrease in farmer income when applying adaptive cropland
43 management measures. For instance, conservation agriculture including the principle of no-till
44 farming contribute to soil erosion management (Chap 6, 6.2. Yet, no-till management can reduce crop
45 yields in some regions, and although this effect is minimised when no-till farming is complemented
46 by the other two principle of conservation agriculture, this could induce a risk to livelihood in
47 vulnerable smallholder farming systems (Pittelkow et al. 2015).

48 Another example is the use of irrigation against water scarcity and drought. During the long lasting
49 drought from 2007–2009 in California, US, farmers adapted by relying on groundwater withdrawal
50 and caused groundwater depletion at unsustainable levels (Christian-Smith et al. 2015). The long term
51 effects of irrigation from groundwater may cause groundwater depletion, land subsidence, aquifer
52 overdraft, and saltwater intrusion (Tularam and Krishna 2009). Therefore, it is expected to increase

1 the vulnerability of coastal aquifers to climate change due to groundwater usage (Ferguson and
2 Gleeson 2012). The long term irrigation practice from groundwater may cause severe combination of
3 potential side effects and consequently irreversible results.

4 **7.2.3.2. Risk associated with land-based mitigation**

5 While historically land use activities have been a net source of GHG emissions, in future decades the
6 land sector will not only need to reduce its emissions, but also to deliver negative emissions through
7 Carbon Dioxide Removal (CDR) to reach the objective of limiting global warming at 2°C or below
8 (Chapter 2 Section 2.5). Although land-based mitigation in itself is a risk-reduction strategy aiming at
9 abating climate change, it also entails risks to humans and ecosystems depending on the type of
10 measures and the scale of deployment. These risks fall broadly into two categories: risk of mitigation
11 failure - due to uncertainties about mitigation potential, potential for sink reversal and moral hazard -
12 and risks arising from adverse side-effects - due to increased competition for land and water
13 resources. This section focuses specifically on bioenergy and BECCS since it is one of the most
14 prominent land-based mitigation strategies in future mitigation scenarios (along with large-scale forest
15 expansion discussed in Cross-Chapter Box 1: Scenarios, in Chapter 1) and it is assessed in Chapter 6
16 as being, at large scales, the only response option with adverse side-effects across all dimensions
17 (adaptation, food security, land degradation and desertification; see 6.4.1).

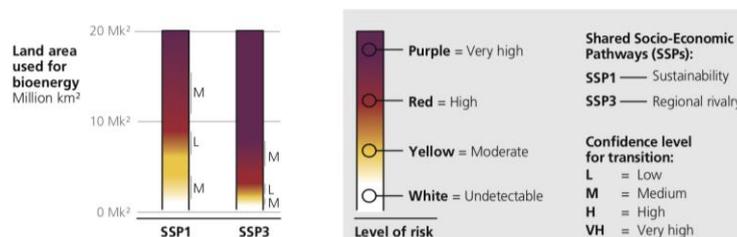
18 *Risk of mitigation failure.* The mitigation potential from bioenergy and BECCS is highly uncertain
19 with estimates ranging from 0.4 to 11.3 GtCO₂e yr⁻¹ for the technical potential while consideration of
20 sustainability constraints suggest an upper end around 5 GtCO₂e yr⁻¹ (Chapter 2, section 2.6). In
21 comparison, IAM-based mitigation pathways compatible with limiting global warming at 1.5°C
22 project bioenergy and BECCS deployment exceeding this range (Chapter 2, Fig. 2.24). There is
23 *medium confidence* that IAMs currently do not reflect the lower end and exceed the upper end of
24 bioenergy and BECCS mitigation potential estimates (Anderson and Peters 2016; Krause et al. 2018;
25 IPCC 2018c), with implications for the risk associated with reliance on bioenergy and BECCS
26 deployment for climate mitigation.

27 In addition, land-based CDR strategies are subject to a risk of carbon sink reversal. This implies a
28 fundamental asymmetry between mitigation achieved through fossil fuel emissions reduction
29 compared to CDR. While carbon in fossil fuel reserves - in the case of avoided fossil fuel emissions -
30 is locked permanently (at least over time scale of several thousand years), carbon sequestered into the
31 terrestrial biosphere – to compensate fossil fuel emissions – is subject to various disturbances in
32 particular from climate change and associated extreme events (Fuss et al. 2018; Dooley and Kartha
33 2018). The probability of sink reversal therefore increases with climate change, implying that the
34 effectiveness of land-based mitigation depends on emission reductions in other sectors and can be
35 sensitive to temperature overshoot (*high confidence*). In the case of bioenergy associated with CCS
36 (BECCS), the issue of the long-term stability of the carbon storage is linked to technical and
37 geological constraints, independent of climate change but presenting risks due to limited knowledge
38 and experience (Chapter 6; Cross-Chapter Box 7: Bioenergy, in Chapter 6).

39 Another factor in the risk of mitigation failure, is the moral hazard associated with CDR technologies.
40 There is *medium evidence and medium agreement* that the promise of future CDR deployment,
41 bioenergy and BECCS in particular, can deter or delay ambitious emission reductions in other sectors
42 (Anderson and Peters 2016; Markusson et al. 2018a; Shue 2018a). The consequences are an increased
43 pressure on land with higher risk of mitigation failure and of temperature overshoot and a transfer of
44 the burden of mitigation and unabated climate change to future generations. Overall, there is therefore
45 *medium evidence and high agreement* that prioritising early decarbonisation with minimal reliance on
46 CDR decreases the risk of mitigation failure and increases intergenerational equity (Geden et al. 2019;
47 Larkin et al. 2018; Markusson et al. 2018b; Shue 2018b).

48 *Risk from adverse side-effects.* At large scales, bioenergy (with or without CCS) is expected to
49 increase competition for land, water resources and nutrients, thus exacerbating the risks of food
50 insecurity, loss of ecosystem services and water scarcity (Chapter 6; Cross-Chapter Box 7: Bioenergy
51 in Chapter 6). Figure 7.3 shows the risk level (from undetectable to very high, aggregating risks of
52 food insecurity, loss of ecosystem services and water scarcity) as a function of the global amount of

1 land (million km²) used for bioenergy, considering second generation bioenergy. Two illustrative
 2 future socio-economic pathways (SSP1 and SSP3; see section 7.2.2 for more details) are depicted, in
 3 SSP3 the competition for land is exacerbated compared to SSP1 due to higher food demand resulting
 4 from larger population growth and higher consumption of meat-based products. The literature used in
 5 this assessment is based on IAM and non-IAM-based studies examining the impact of bioenergy crop
 6 deployment on various indicators, including food security (food prices or population at risk of hunger
 7 with explicit consideration of exposure and vulnerability), SDGs, ecosystem losses, transgression of
 8 various planetary boundaries and water consumption (see supplementary material). Since most of the
 9 assessed literature is centered around 2050 prevailing demographic and economic conditions for this
 10 year are used for the risk estimate. An aggregated risk metric including risks of food insecurity, loss
 11 of ecosystem services and water scarcity is used because there is no unique relationship between
 12 bioenergy deployment and the risk outcome for a single system. For instance, bioenergy deployment
 13 can be implemented in such a way that food security is prioritised at the expense of natural
 14 ecosystems, while the same scale of bioenergy deployment implemented with ecosystem safeguards
 15 would lead to a fundamentally different outcome in terms of food security (Boysen et al. 2017a).
 16 Considered as a combined risk, however, the possibility of a negative outcome on either food security,
 17 ecosystems or both can be assessed with less ambiguity and independently of possible implementation
 18 choices.
 19



20
21

22 **Figure 7.3: Risks associated with bioenergy crop deployment as a land-based mitigation strategy under**
 23 **two SSPs (SSP1 and SSP3). The assessment is based on literature investigating the consequences of**
 24 **bioenergy expansion for food security, ecosystem loss and water scarcity. These risk indicators were**
 25 **aggregated as a single risk metric in the figure. In this context, very high risk indicates that important**
 26 **adverse consequences are expected for all these indicators (more than 100 million people at risk of**
 27 **hunger, major ecosystem losses and severe water scarcity issues). The climate scenario considered is a**
 28 **mitigation scenario consistent with limiting global warming at 2°C (RCP2.6), however some studies**
 29 **considering other scenarios (e.g., no climate change) were considered in the expert judgement as well as**
 30 **results from other SSPs (e.g., SSP2). The literature supporting the assessment is provided in Table SM7.3.**

31 In SSP1, there is *medium confidence* that 1 to 4 million km² can be dedicated to bioenergy production
 32 without significant risks to food security, ecosystem services and water scarcity. At these scales of
 33 deployment, bioenergy and BECCS could have co-benefits for instance by contributing to restoration
 34 of degraded land and soils (Cross-Chapter Box 7: Bioenergy and BECCS in Chapter 6). Although
 35 currently degraded soils (up to 20 million km²) represent a large amount of potentially available land
 36 (Boysen et al. 2017a), trade-offs would occur already at smaller scale due to fertiliser and water use
 37 (Hejazi et al. 2014; Humpenöder et al. 2017; Heck et al. 2018a; Boysen et al. 2017b). There is *low*
 38 *confidence* that the transition from moderate to high risk is in the range 6-8.7 million km². In SSP1,
 39 (Humpenöder et al. 2017) found no important impacts on sustainability indicators at a level of 6.7
 40 million km², while (Heck et al. 2018b) note that several planetary boundaries (biosphere integrity;
 41 land-system change; biogeochemical flows; freshwater use) would be exceeded above 8.7 million
 42 km². There is *very high confidence* that all the risk transitions occur at lower bioenergy levels in
 43 SSP3, implying higher risks associated with bioenergy deployment, due to the higher competition for
 44 land in this pathway. In SSP3, land-based mitigation is therefore strongly limited by sustainability
 45 constraints such that moderate risk occur already between 0.5 and 1.5 million km² (*medium*

1 *confidence*). There is *medium confidence* that a bioenergy footprint beyond 4 to 8 million km² would
2 entail very high risk with transgression of most planetary boundaries (Heck et al. 2018b), strong
3 decline in sustainability indicators (Humpenöder et al. 2017) and increase in the population at risk of
4 hunger well above 100 million (Fujimori et al. 2018a; Hasegawa et al. 2018b).

6 **7.2.4. Risks arising from Hazard, Exposure, and Vulnerability**

7 Table 7.1 shows hazards from land-climate-society interactions identified in previous chapters, or in
8 other IPCC reports (with supplementary hazards appearing in the Appendix); the regions that are
9 exposed or will be exposed to these hazards; components of the land-climate systems and societies
10 that are vulnerable to the hazard; the risk associated with these impacts and the available indicative
11 policy responses. The last column shows representative supporting literature.

12 Included are forest dieback, extreme events in multiple economic and agricultural regimes (also see
13 7.2.2.1, 7.2.2.2), disruption in flow regimes in river systems, climate change mitigation impacts (also
14 see 7.2.3.2), competition for land (plastic substitution by cellulose, charcoal production), land
15 degradation and desertification (also see 7.2.2.8), loss of carbon sinks, permafrost destabilisation (also
16 see 7.2.2.7), and stranded assets (also see 7.3.4). Other hazards such as from failure of carbon storage,
17 renewable energy impacts on land use, wild-fire in forest-urban transition context, extreme events
18 effects on cultural heritage and urban air pollution from surrounding land-use are covered in Table 7.1
19 extension in the appendix as well in 7.5.6.

20

21

1
2
3
4
5
6

Table 7.1 Characterising land-climate risk and indicative policy responses. Table shows hazards from land-climate-society interactions identified in previous chapters or in *other* IPCC reports; the regions that are exposed or will be exposed to these hazards; components of the land-climate systems and societies that are vulnerable to the hazard; the risk associated with these impacts and the available policy responses and response options from Chapter 6. The last column shows representative supporting literature

| Land-Climate-Society interaction Hazard | Exposure | Vulnerability | Risk | Policy Response (Indicative) | References |
|---|--------------------------------------|---|---|--|---|
| Forest dieback | Widespread across biomes and regions | Marginalised Population with insecure land tenure | <ul style="list-style-type: none"> • Loss of forest-based livelihoods • Loss of identity | <ul style="list-style-type: none"> • Land rights • Community based conservation • Enhanced political enfranchisement • Manager-scientist partnerships for adaptation silviculture | (Allen et al. 2010; McDowell and Allen 2015; Sunderlin et al. 2017; Belcher et al. 2005; Soizic et al 2013)(Nagel et al. 2017) |
| | | Endangered species and ecosystems | <ul style="list-style-type: none"> • Extinction • Loss of ecosystem services • Cultural loss | <ul style="list-style-type: none"> • Effective enforcement of protected areas and curbs on illegal trade • Ecosystem Restoration • Protection of indigenous people | (Bailis et al. 2015; Cameron et al. 2016) |
| Extreme events in multiple economic and agricultural regimes | Global | <ul style="list-style-type: none"> • Food importing countries • Low income indebtedness • Net food buyer | <ul style="list-style-type: none"> • Conflict • Migration • Food inflation • Loss of life • Disease, malnutrition • Farmer suicides | <ul style="list-style-type: none"> • Insurance • Social Protection encouraging diversity of sources • Climate smart agriculture • Land rights and tenure • Adaptive Public Distribution Systems | (Fraser et al. 2005; Schmidhuber and Tubiello 2007; Lipper et al. 2014a; Lunt et al. 2016; Tigchelaar et al. 2018; Casellas Connors and Janetos 2016) |

| Land-Climate-Society interaction Hazard | Exposure | Vulnerability | Risk | Policy Response (Indicative) | References |
|--|--|--|---|---|---|
| Disruption of flow regimes in river systems | 1.5 billion people, Regional (e.g., South Asia, Australia) Aral sea and others | <ul style="list-style-type: none"> • Water intensive agriculture • Fresh-water, estuarine and near coastal ecosystems • Fishers • Endangered species and ecosystems | <ul style="list-style-type: none"> • Loss of livelihoods and identity • Migration • Indebtedness | <ul style="list-style-type: none"> • Build alternative scenarios for economies and livelihoods based on non-consumptive use (e.g., wild capture fisheries) • Define and maintain ecological flows in rivers for target species and ecosystem services • Experiment with alternative less water consuming crops and water management strategies • Redefine SDGs to include fresh-water ecosystems or adopt alternative metrics of sustainability Based on Nature Contributions to People (NCP) | (Craig 2010; Di Baldassarre et al. 2013; Verma et al. 2009; Ghosh et al. 2016; Higgins et al. 2018;) (Hall et al. 2013; Youn et al. 2014) |
| Depletion/ exhaustion of ground-water | Wide-spread across semi-arid and humid biomes India, China and the United States Small Islands | <ul style="list-style-type: none"> • Farmers, drinking water supply • Irrigation • See forest note above • Agricultural production • Urban sustainability (Phoenix, US) | <ul style="list-style-type: none"> • Food insecurity • Water insecurity • Distress migration • Conflict • Disease • Inundation of coastal regions, estuaries and deltas | <ul style="list-style-type: none"> • Monitoring of emerging ground-water-climate linkages • Adaptation strategies that reduce dependence on deep ground water • Regulation of ground-water use • Shift to less water- | (Wada et al. 2010; Rodell et al. 2009; Taylor et al. 2013; Aeschbach-Hertig and Gleeson 2012) |

| Land-Climate-Society interaction Hazard | Exposure | Vulnerability | Risk | Policy Response (Indicative) | References |
|--|---|--|--|---|--|
| Climate change Mitigation impacts | Across various biomes especially semi-arid and aquatic where renewable energy projects (solar, biomass, wind and small hydro) are sited | <ul style="list-style-type: none"> • Reduction in dry-season river flows • Sea level rise • Fishers and pastoralists • Farmers • Endangered range restricted species and ecosystems | <ul style="list-style-type: none"> • Extinction of species • Downstream loss of ecosystem services Loss of livelihoods and identity of fisher/pastoralist communities • Loss of regional food security | <p>intensive rain fed crops and pasture</p> <ul style="list-style-type: none"> • Conjunctive use of surface and ground-water • Avoidance and informed siting in priority basins • Mitigation of impacts • Certification | (Zomer et al. 2008; Nyong et al. 2007; Pielke et al. 2002; Schmidhuber and Tubiello 2007; Jumani et al. 2017; Eldridge et al. 2011; Bryan et al. 2010; Scarlat and Dallemand 2011) |
| Competition for land substitution by e.g., Plastic cellulose, Charcoal production | Peri-urban and rural areas in developing countries | <ul style="list-style-type: none"> • Rural landscapes; farmers; charcoal suppliers; small businesses | <ul style="list-style-type: none"> • Land degradation; loss of ecosystem services; GHG emissions; lower adaptive capacity | <ul style="list-style-type: none"> • Sustainability certification; producer permits; subsidies for efficient kilns | (Woollen et al. 2016; Kiruki et al. 2017a) |
| Land degradation and desertification | Arid, Semi-arid and sub-humid regions | <ul style="list-style-type: none"> • Farmers • Pastoralists • Biodiversity | <ul style="list-style-type: none"> • Food insecurity • Drought • Migration • Loss of agro and wild biodiversity | <ul style="list-style-type: none"> • Restoration of ecosystems and management of invasive species • Climate smart agriculture and | (Fleskens, Luuk, Stringer 2014; Lambin et al. 2001; Cowie et al. 2018a; Few and Tebboth 2018; Sandstrom and |

| Land-Climate-Society interaction Hazard | Exposure | Vulnerability | Risk | Policy Response (Indicative) | References |
|---|--|---|--|--|---|
| Loss of carbon sinks | Wide-spread across biomes and regions | <ul style="list-style-type: none"> • Tropical forests • Boreal soils | <ul style="list-style-type: none"> • Feed-back to global and regional climate change | livestock management <ul style="list-style-type: none"> • Managing economic impacts of global and local drivers • Changes in relief and rehabilitation policies • Land degradation neutrality • Conservation prioritisation of tropical forests • Afforestation | Juhola 2017) (Barnett et al. 2005; Tribbia and Moser 2008) |
| Permafrost destabilisation | Arctic and Sub-Arctic regions | <ul style="list-style-type: none"> • Soils • Indigenous communities • Biodiversity | <ul style="list-style-type: none"> • Enhanced GHG emissions | <ul style="list-style-type: none"> • Enhanced carbon uptake from novel ecosystem after thaw • Adapt to emerging wetlands | (Schuur et al. 2015) |
| Stranded assets | Economies transitioning to low carbon pathways Oil economies Coastal regions facing inundation | Coal based power Oil refineries Plastic industry Large dams Coastal infrastructure | <ul style="list-style-type: none"> • Disruption of regional economies and conflict • Unemployment • Push-back against renewable energy • Migration | <ul style="list-style-type: none"> • Insurance and tax cuts • Long-term power purchase agreements • Economic and technical support for transitioning economies • transforming oil wealth into renewable energy leadership • Redevelopment using adaptation • OPEC investment in information sharing for transition | (Farfan and Breyer 2017; Ansar et al. 2013; Van de Graaf 2017; Trieb et al. 2011) |

1

2

3 **7.3. Consequences of climate – land change for human well-being and** 4 **sustainable development**

5 To further explore what is at stake for human systems, this section assesses literature about potential
6 consequences of climate and land change for human well-being and ecosystems upon which humans
7 depend. Risks described in 7.2 have significant social, spiritual, and economic ramifications for
8 societies across the world and this section explores potential implications of the risks outlined above
9 to food security, livelihood systems, migration, ecosystems, species, infectious disease, and
10 communities and infrastructure. Because food and livelihood systems are deeply tied to one another,
11 combinations of climate and land change could pose higher present risks to humans and ecosystems
12 than examination of individual elements alone might suggest.

13 **7.3.1. What is at stake for food security?**

14 This section examines risks to food security when access to food is jeopardised by yield shortfall and
15 instability related to climate stressors. Past assessments of climate change impacts have sometimes
16 assumed that when grain and food yields in one area of the world are lower than expected, world trade
17 can redistribute food adequately to ensure food security. There is *medium confidence* that severe and
18 spatially extensive climatic stressors pose *high risk* to stability of and access to food for large numbers
19 of people across the world.

20 The 2007–2008, and 2010–2011 droughts in several regions of the world resulted in crop yield
21 decline that in turn led some governments to protect their domestic grain supplies rather than
22 engaging in free trade to offset food shortfalls in other areas of the world. These responses cascaded
23 and strongly affected regional and global food prices. Simultaneous crop yield impacts combined with
24 trade impacts have proven to play a larger and more pervasive role in global food crises than
25 previously thought (Sternberg 2012, 2017; Bellemare 2015) (Chatzopoulos et al. 2019). There is *high*
26 *confidence* that regional climate extremes already have significant negative domestic and international
27 economic impacts (Chatzopoulos et al. 2019).

28 **7.3.2. Risks to where and how people live: Livelihood systems and migration**

29 There is *high confidence* that climate- and land change interact with social, economic, political, and
30 demographic factors that affect how well and where people live (Sudmeier-Rieux et al. 2017;
31 Government Office for Science 2011; Laczko and Piguat 2014; Bohra-Mishra and Massey 2011;
32 Raleigh et al. 2015; Warner and Afifi 2011; Hugo 2011; Warner et al. 2012). There is *high evidence*
33 *and high agreement* that people move to manage risks and seek opportunities for their safety and
34 livelihoods, recognising that people respond to climatic change and land-related factors in tandem
35 with other variables (Hendrix and Salehyan 2012; Lashley and Warner 2015; van der Geest and
36 Warner 2014; Roudier et al. 2014; Warner and Afifi 2014)(McLeman 2013; Kaenzig and Piguat 2014;
37 Internal Displacement Monitoring Center 2017; Warner 2018; Cohen and Bradley 2010; Thomas and
38 Benjamin 2017). People move towards areas offering safety and livelihoods such as in rapidly
39 growing settlements in coastal zones (Black et al. 2013; Challinor et al. 2017; Adger et al. 2013);
40 burgeoning urban areas also face changing exposure to combinations of storm surges and sea level
41 rise, coastal erosion and soil and water salinisation, and land subsidence (Geisler and Currens 2017;
42 Maldonado et al. 2014; Bronen and Chapin 2013).

43 There is *medium confidence* that livelihood-related migration can accelerate in the short to medium
44 term when weather dependent livelihood systems deteriorate in relation to changes in precipitation,
45 changes in ecosystems, and land degradation and desertification (Abid et al. 2016)(Scheffran et al.
46 2012; Fussell et al. 2014; Bettini and Gioli 2016; Reyer et al. 2017)(Warner and Afifi 2014)(Handmer
47 et al. 2012; Nawrotzki and Bakhtsiyarava 2017; Nawrotzki et al. 2016; Steffen et al. 2015; Black et al.

1 2013). Slow onset climate impacts and risks can exacerbate or otherwise interact with social conflict
2 corresponding with movement at larger scales (see Section 7.2.3.2) and long term deterioration in
3 habitability of regions could trigger spatial population shifts (Denton et al. 2014).

4 There is *medium evidence* and *medium agreement* that climatic stressors can worsen the complex
5 negative impacts of strife and conflict (Schleussner et al. 2016; Barnett and Palutikof 2014; Scheffran
6 et al. 2012). Climate change and human mobility could be a factor that heightens tensions over scarce
7 strategic resources, a further destabilising influence in fragile states experiencing socio-economic and
8 political unrest (Carleton and Hsiang 2016a). Conflict and changes in weather patterns can worsen
9 conditions for people working in rain fed agriculture or subsistence farming, interrupting production
10 systems, degrading land and vegetation further (Papaioannou 2016; Adano and Daudi 2012). In recent
11 decades, droughts and other climatic stressors have compounded livelihood pressures in areas already
12 torn by strife (Tessler et al. 2015; Raleigh et al. 2015), such as in the Horn of Africa. Seizing of
13 agricultural land by competing factions, preventing food distribution in times of shortage have in this
14 region and others contributed to a triad of food insecurity, humanitarian need, and large movements of
15 people (Theisen et al. 2011; Mohammed et al. 2018; Ayeb-Karlsson et al. 2016; von Uexkull et al.
16 2016; Gleick 2014; Maystadt and Ecker 2014). People fleeing complex situations may return if
17 peaceful conditions can be established. Climate change and climate change induced development
18 responses in countries and regions are likely to exacerbate tensions over water and land its impact on
19 agriculture, fisheries, livestock and drinking water downstream. Shared pastoral landscapes used by
20 disadvantaged or otherwise vulnerable communities are particularly impacted by conflicts that are
21 likely to become more severe under future climate change (Salehyan and Hendrix 2014; Hendrix and
22 Salehyan 2012). Extreme events could considerably enhance these risks, in particular long-term
23 drying trends (Kelley et al. 2015; Cutter et al. 2012a). There is *medium evidence* and *medium*
24 *agreement* that governance is key in magnifying or moderating climate change impact and conflict
25 (Bonatti et al. 2016).

26 There is *low evidence and medium agreement* that longer-term deterioration in the habitability of
27 regions could trigger spatial population shifts (Seto 2011). Heat waves, rising sea levels that salinise
28 and inundate coastal and low-lying aquifers and soils, desertification, loss of geologic sources of
29 water such as glaciers and freshwater aquifers could affect many regions of the world and put life-
30 sustaining ecosystems under pressure to support human populations (Flahaux and De Haas 2016;
31 Chambwera et al. 2015; Tierney et al. 2015; Lilleør and Van den Broeck 2011).

32 **7.3.3. Risks to humans from disrupted ecosystems and species**

33 **Risks of loss of biodiversity and ecosystem services**

34 Climate change poses significant threat to species survival, and to maintaining biodiversity and
35 ecosystem services. Climate change reduces the functionality, stability, and adaptability of
36 ecosystems (Pecl et al. 2017). For example, drought affects cropland and forest productivity and
37 reduces associated harvests (provisioning services). In addition, extreme changes in precipitation
38 may reduce the capacity of forests to provide stability for groundwater (regulation and maintenance
39 services). Prolonged periods of high temperature may cause widespread death of trees in tropical
40 mountains, boreal and tundra forests, impacting diverse ecosystem services including impacting
41 aesthetic and cultural services (Verbyla 2011; Chapin et al. 2010; Krishnaswamy et al. 2014).
42 According to the Millennium Ecosystem Assessment (Millennium Ecosystem Assessment 2005),
43 climate change is likely to become one of the most significant drivers of biodiversity loss by the end
44 of the century.

45 There is *high confidence* that climate change already poses a moderate risk to biodiversity, and is
46 projected to become a progressively widespread and high risk in the coming decades; loss of Arctic
47 sea ice threatens biodiversity across an entire biome and beyond; the related pressure of ocean
48 acidification, resulting from higher concentrations of carbon dioxide in the atmosphere, is also already
49 being observed (UNEP 2009). There is ample evidence that climate change and land change
50 negatively affects biodiversity across wide spatial scales. Although there is relatively *limited evidence*
51 of current extinctions caused by climate change, studies suggest that climate change could surpass
52 habitat destruction as the greatest global threat to biodiversity over the next several decades (Pereira et

1 al. 2010). However, the multiplicity of approaches and the resulting variability in projections make it
2 difficult to get a clear picture of the future of biodiversity under different scenarios of global climatic
3 change (Pereira et al. 2010) . Biodiversity is also severely impacted by climate change induced land
4 degradation and ecosystem transformation (Pecl et al. 2017). This may impact humans directly and
5 indirectly through cascading impacts on ecosystem function and ecosystem services (Millennium
6 Assessment 2005). Climate change related human migration is likely to impact biodiversity as people
7 move into and contribute to land stress in biodiversity hotspots now and in the future; and as humans
8 concurrently move into areas where biodiversity is also migrating to adapt to climate change
9 (Oglethorpe et al. 2007).

10 **Climate and land change increases risk to respiratory and infectious disease**

11 In addition to risks related to nutrition articulated in Figure 7.1, human health can be affected by
12 climate change through extreme heat and cold, changes in infectious diseases, extreme events, and
13 land cover and land use (Hasegawa et al. 2016; Ryan et al. 2015; Terrazas et al. 2015; Kweka et al.
14 2016; Yamana et al. 2016). Evidence indicates that action to prevent the health impacts of climate
15 change could provide substantial economic benefits (Martinez et al. 2015; Watts et al. 2015).

16 Climate change exacerbates air pollution with increasing UV and ozone concentration. It has negative
17 impacts on human health and increases mortality rate especially in urban region (Silva et al. 2016,
18 2013; Lelieveld et al. 2013; Whitmee et al. 2015; Anenberg et al. 2010). In the Amazon, research
19 shows that deforestation (both net loss and fragmentation) will increase malaria, where vectors are
20 expected to increase their home range (Alimi et al. 2015; Ren et al. 2016), confounded with multiple
21 factors, such as social-economic conditions and immunity (Tucker Lima et al. 2017; Barros and
22 Honório 2015). Deforestation has been shown to enhance the survival and development of major
23 malaria vectors (Wang et al. 2016). The WHO estimates 60,091 additional deaths for climate change
24 induced malaria for the year 2030 and 32,695 for 2050 (World Health Organization 2014).

25 Human encroachment on animal habitat in combination with the bushmeat trade in Central African
26 countries has contributed to the increased incidence of zoonotic (i.e., animal-derived) diseases in
27 human populations, including Ebola virus epidemic (Alexander et al. 2015a; Nkengasong and
28 Onyebujoh 2018). The composition and density of zoonotic reservoir populations, such as rodents, is
29 also influenced by land-use and climate change (*high confidence*) (Young et al. 2017a). The bushmeat
30 trade in many regions of central and west African forests (particularly in relation to chimpanzee and
31 gorilla populations) elevates the risk of ebola by increasing human-animal contact (Harrod 2015).

32 **7.3.4. Risks to Communities and Infrastructure**

33 There is *high confidence* that policies and institutions which accentuate vicious cycles of poverty and
34 ill-health, land degradation and greenhouse gas emissions undermine stability and are barriers to
35 achieving climate resilient sustainable development. There is *high confidence* that change in climate
36 and land pose high periodic and sustained risk to the very young, those living in poverty, and ageing
37 populations. Older people are particularly exposed due to more restricted access to resources, changes
38 in physiology, and decreased mobility resulting from age which may limit adaptive capacity of
39 individuals and populations as a whole (Filiberto et al. 2010).

40 Combinations of food insecurity, livelihood loss related to degrading soils and ecosystem change, or
41 other factors that diminish the habitability of where people live disrupt social fabric and are currently
42 detected in most regions of the world (Carleton and Hsiang 2016b) There is *high confidence* that
43 coastal flooding and degradation already poses widespread and rising future risk to infrastructure
44 value and stranded infrastructure, as well as livelihoods made possible by urban infrastructure
45 (Radhakrishnan et al. 2017; Pathirana, A., Radhakrishnan, M., Quan 2018; Pathirana, A.,
46 Radhakrishnan, M., Ashley 2018; Radhakrishnan, M., Nguyen, H., Gersonius 2018; EEA 2016;
47 Pelling and Wisner 2012; Oke et al. 2017; Parnell and Walawege 2011; Uzun and Cete 2004; Melvin
48 et al. 2017).

49 There is *high evidence and high agreement* that climate and land change pose high risk to
50 communities and interdependent infrastructure systems including electric power, and transportation
51 are highly vulnerable and interdependent (Below et al. 2012; Adger et al. 2013; Pathirana, A.,

1 Radhakrishnan, M., Quan 2018)(Conway and Schipper 2011; Caney 2014; Chung Tiam Fook 2017;
2 Pathirana, A., Radhakrishnan, M., Quan 2018). These systems are exposed to disruption from severe
3 climate events such as weather-related power interruptions lasting for hours to days (Panteli and
4 Mancarella 2015). Increased magnitude and frequency of high winds, ice storms, hurricanes and heat
5 waves have caused widespread damage to power infrastructure and have caused severe outages,
6 affecting significant numbers of customers in urban and rural areas (Abi-Samra and Malcolm 2011).

7 Increasing populations, enhanced per capita water use, climate change, and allocations for water
8 conservation are potential threats to adequate water availability. As climate change produces
9 variations in rainfall, these challenges will intensify, evidenced by severe water shortages in recent
10 years in Capetown, Los Angeles, Rio de Janeiro among others (Watts et al. 2018; Majumder 2015;
11 Ashoori et al. 2015; Mini et al. 2015; Otto et al. 2015)(Cross-Chapter Box 5: Case study on policy
12 responses to drought in Chapter 3)(Ranatunga et al. 2014)(Ray and Shaw 2016; Gopakumar 2014).

14 **Cross-chapter Box 10: Economic dimensions of climate change** 15 **and land**

16 Koko Warner (The United States of America), Aziz Elbehri (Morocco), Marta Guadalupe Rivera
17 Ferre (Spain), Alisher Mirzabaev (Germany/Uzbekistan), Lindsay Stringer (United Kingdom), Anita
18 Wreford (New Zealand)

19
20 Sustainable land management (SLM) makes strong social and economic sense. Early action in
21 implementing SLM for climate change adaptation and mitigation provides distinct societal
22 advantages. Understanding the full scope of what is at stake from climate change presents challenges
23 because of inadequate accounting of the degree and scale at which climate change and land
24 interactions impact society, and the importance society places on those impacts (Santos et al.
25 2016)(7.2.2, 5.3.1, 5.3.2, 4.1). The consequences of inaction and delay bring significant risks
26 including irreversible change and loss in land ecosystem services, including food security, with
27 potentially substantial economic damage to many countries in many regions of the world (*high*
28 *confidence*).

29
30 This cross-chapter box brings together the salient economic concepts underpinning the assessments of
31 sustainable land management and mitigation options presented in this report. Four critical concepts
32 are required to help assess the social and economic implications of land-based climate action:

- 33 i. value to society;
- 34 ii. damages from climate and land-induced interventions on land ecosystems;
- 35 iii. costs of action and inaction;
- 36 iv. decision-making under uncertainty.

37 38 (i) **Value to society**

39 Healthy functioning land and ecosystems are essential for human health, food and livelihood security.
40 Land derives its value to humans from being both a finite resource and vital for life, providing vital
41 ecosystem services from water recycling, food, feed, fuel, biodiversity and carbon storage and
42 sequestration.

43
44 Many of these ecosystem services may be difficult to estimate in monetary terms, including when
45 they hold high symbolic value, linked to ancestral history, or traditional and indigenous knowledge
46 systems (Boillat and Berkes 2013). Such incommensurable values of land are core to social
47 cohesion— social norms and institutions, trust that enables all interactions, and sense of community.

48 49 (ii) **Damages from climate and land-induced interventions on land ecosystems;**

50 Values of many land-based ecosystem services and their potential loss under land-climate change
51 interaction can be considerable: the global value of ecosystem services was valued in 2011 at USD

1 125 trillion per year and the annual loss due to land use change was between USD 4.3 to 20.2 trillion
2 per year from 2007 (Costanza et al. 2014; Rockström et al. 2009). The annual costs of land
3 degradation are estimated to be about USD 231 billion per year or about 0.41% of the global GDP of
4 USD 56.49 trillion in 2007 (Nkonya et al. 2016) (4.4.1, 4.4.2).

5
6 Studies show increasingly negative effects on GDP from damage and loss to land-based values and
7 service as global mean temperatures increase, although the impact varies across regions (Kompas et
8 al. 2018).

9 10 **(iii) Costs of action and inaction**

11 Evidence suggests that the cost of inaction in mitigation and adaptation, and land use, exceeds the cost
12 of interventions in both individual countries, regions, and worldwide (Nkonya et al. 2016). Continued
13 inaction reduces the future policy option space, dampens economic growth and increases the
14 challenges of mitigation as well as adaptation (Moore and Diaz 2015)(Luderer et al. 2013). The cost
15 of reducing emissions is estimated to be considerably less than the costs of the damages at all levels
16 (Kainuma et al. 2013; Moran 2011; Sánchez and Maseda 2016).

17
18 The costs of adapting to climate impacts are also projected to be substantial, although evidence is
19 limited (summarised in Chambwera et al. 2014a). Estimates range from USD 9 to 166 billion per year
20 at various scales and types of adaptation, from capacity building to specific projects (Fankhauser
21 2017). Inadequate literature exists on the costs of adaptation in the agriculture or land-based sectors
22 (Wreford and Renwick 2012) due to lack of baselines, uncertainty around biological relationships and
23 inherent uncertainty about anticipated avoided damage estimates, but economic appraisal of actions to
24 maintain the functions of the natural environment and land sector generate positive net present values
25 (Adaptation Sub-committee 2013).

26
27 Preventing land degradation from occurring is considered more cost-effective in the long term
28 compared to the magnitude of resources required to restore already degraded land (Cowie et al.
29 2018a) (3.6.1). Evidence from drylands shows that each US dollar invested in land restoration
30 provides between 3 and 6 in social returns over a 30 year period, using a discount rate between 2.5
31 and 10% (Nkonya et al. 2016). SLM practices reverse or minimise economic losses of land
32 degradation, estimated at between USD 6.3 and 10.6 trillion annually, (ELD Initiative 2015) more
33 than five times the entire value of agriculture in the market economy (Costanza et al. 2014; Fischer et
34 al. 2017; Sandifer et al. 2015; Dasgupta et al. 2013) (3.7.5).

35
36 Across other areas such as food security, disaster mitigation and risk reduction, humanitarian
37 response, and healthy diet (malnutrition as well as disease), early action generates economic benefits
38 greater than costs (*high evidence, high agreement*) (Fankhauser 2017; Wilkinson et al. 2018; Venton
39 2018; Venton et al. 2012) (Clarvis et al. 2015)(Nugent et al. 2018) (Watts et al. 2018) (Bertram et al.
40 2018)(6.3, 6.4).

41 42 **(iv) Decision-making under uncertainty**

43 Given that significant uncertainty exists regarding the future impacts of climate change, effective
44 decisions must be made under unavoidable uncertainty (Jones et al., 2014).

45 Approaches that allow for decision-making under uncertainty are continually evolving (see 7.5). An
46 emerging trend is towards new frameworks that will enable multiple decision makers with multiple
47 objectives to explore the trade-offs between potentially conflicting preferences to identify strategies
48 that are robust to deep uncertainties (Singh et al. 2015; Driscoll et al. 2016; Araujo Enciso et al. 2016;
49 Herman et al. 2014; Pérez et al. 2016; Girard et al. 2015; Haasnoot et al. 2018; Roelich and Gieseckam
50 2019).

51 52 **Valuation of benefits and damages and costing interventions: Measurement issues**

53 Cost appraisal tools for climate adaptation are many and their suitability depends on the context
54 (7.5.2.2). Cost-benefit analysis (CBA) and cost-effectiveness analysis (CEA) are commonly applied,
55 especially for current climate variability situations. However, these tools are not without criticism and

1 their limitations have been observed in the literature (see Rogelj et al. 2018). In general measuring
 2 costs and providing valuation are influenced by four conditions: measurement and valuation; the time
 3 dimension; externalities; and aggregate versus marginal costs:

4
 5 **Measurement and value issues**

6 Ecosystem services that are not traded in the market fall outside the formal or market-based valuation
 7 and their value is thus either not accounted for or underestimated in both private and public decisions
 8 (Atkinson et al. 2018). Environmental valuation literature uses a range of techniques to assign
 9 monetary values to environmental outcomes where no market exists (Atkinson et al. 2018) (Dallimer
 10 et al. 2018), but some values remain inestimable. For some indigenous cultures and peoples, land is
 11 not considered something that can be sold and bought, so economic valuations are not meaningful
 12 even as proxy approaches (Boillat and Berkes 2013)(Kumpula et al. 2011; Pert et al. 2015; Xu et al.
 13 2005).

14
 15 While a rigorous CBA is broader than a purely financial tool and can capture non-market values
 16 where they exist, it can prioritise certain values over others (such as profit maximisation for owners,
 17 efficiency from the perspective of supply chain processes, and judgements about which parties bear
 18 the costs). Careful consideration of whose perspectives are considered when undertaking a CBA and
 19 the limitations of these methods for policy interventions.

20
 21 **Time dimension (short vs long term) and the issue of discount rates**

22
 23 Economics uses a mechanism to convert future values to present day values known as discounting, or
 24 the pure rate of time preference. Discount rates are increasingly being chosen to reflect concerns
 25 about intergenerational equity, and some countries (e.g., the UK and France) apply a declining
 26 discount rate for long term public projects. The choice of discount rate has important implications for
 27 policy evaluation (Anthoff, Tol, & Yohe, 2010; Arrow et al., 2014; Baral, Keenan, Sharma, Stork, &
 28 Kasel, 2014; Dasgupta et al., 2013; Lontzek, Cai, Judd, & Lenton, 2015; Sorokin et al., 2015; van den
 29 Bergh & Botzen, 2014)(*high evidence, high agreement*). Stern (Stern 2007), for example, used a much
 30 lower discount rate (giving almost equal weight to future generations) than the mainstream authors
 31 (e.g., Nordhaus) and obtained much higher estimates of the damage of climate change.

32
 33 **Positive and negative externalities (consequences and impacts not accounted for in
 34 market economy),**

35 All land use generates externalities (unaccounted for side-effects of an activity). Examples include
 36 loss of ecosystem services (e.g., reduced pollinators; soil erosion, increased water pollution,
 37 nitrification etc.). Positive externalities include sequestration of CO₂ and improved soil water
 38 filtration from afforestation. Externalities can also be social (e.g., displacement and migration) and
 39 economic (e.g., loss of productive land). In the context of climate change and land, the major
 40 externality is the AFOLU sourced emissions of GHGs. Examples of mechanisms to internalise
 41 externalities are discussed in 7.5.

42
 43 **Aggregate versus marginal costs**

44
 45 Costs of climate change are often referred to through the marginal measure of the Social Cost of
 46 Carbon (SCC), which measures the total net damages of an extra metric ton of CO₂ emissions due to
 47 the associated climate change (Nordhaus 2014). The SCC can be used to determine a carbon price, but
 48 SCC depends on discount rate assumptions and may neglect processes including large losses of
 49 biodiversity, political instability, violent conflicts, large-scale migration flows, and the effects of
 50 climate change on the development of economies (Stern 2013; Pezzey 2019).

51
 52 At the sectoral level, marginal abatement cost (MAC) curves are widely used for the assessment of
 53 costs related to CO₂ or GHG emissions reduction. MAC measures the cost of reducing one more GHG
 54 unit and MAC curves are either expert-based or model-derived and offer a range of approaches and

1 assumptions on discount rates or available abatement technologies (Moran 2011).

3 **7.3.4.1. Windows of Opportunity**

4 Windows of opportunity are important learning moments wherein an event or disturbance in relation
5 to land, climate, and food security triggers responsive social, political, policy change (*medium*
6 *agreement*). Policies play an important role in windows of opportunity and are important in relation to
7 managing risks of desertification, soil degradation, food insecurity, and supporting response options
8 for sustainable land management (Chapter 6) (*high agreement*) (Kivimaa and Kern 2016; Gupta et al.
9 2013b; Cosens et al. 2017b; Darnhofer 2014; Duru et al. 2015).

10 A wide range of events or disturbances may initiate windows of opportunity ranging from climatic
11 events and disasters, recognition of a state of land degradation, an ecological social or political crisis,
12 and a triggered regulatory burden or opportunity. Recognition of a degraded system such as land
13 degradation and desertification (Chapters 3 and 4) and associated ecosystem feedbacks, allows for
14 strategies, response options and policies to address the degraded state (Nyström et al. 2012). Climate
15 related disasters (flood, droughts etc.) and crisis may trigger latent local adaptive capacities leading to
16 systemic equitable improvement (McSweeney and Coomes 2011), or novel and innovative
17 recombining of sources of experience and knowledge, allowing navigation to transformative social
18 ecological transitions (Folke et al. 2010). The occurrence of a series of punctuated crisis such as
19 floods or droughts, qualify as windows of opportunity when they enhance society's capacity to adapt
20 over the long term (Pahl-Wostl et al. 2013). A disturbance from an ecological, social, or political
21 crisis may be sufficient to trigger the emergence of new approaches to governance wherein there is a
22 change in the rules of the social world such as informal agreements surrounding human activities or
23 formal rules of public policies (Olsson et al. 2006; Biggs et al. 2017) (See 7.6). A combination of
24 socio-ecological changes may provide windows of opportunity for a socio-technical niche to be
25 adopted on a greater scale transforming practices towards sustainable land management such as
26 biodiversity based agriculture (Darnhofer 2014; Duru et al. 2015).

27 Policy may also create windows of opportunity. A disturbance may cause inconvenience, including
28 high costs of compliance with environmental regulations, thereby initiating a change of behaviour
29 (Cosens et al. 2017a). In a similar vein, multiple regulatory requirements existing at the time of a
30 disturbance may result in emergent processes and novel solutions in order to correct for piecemeal
31 regulatory compliance (Cosens et al. 2017a). Lastly, windows of opportunity can be created by policy
32 mixes or portfolio that provide for creative destruction of old social processes and thereby encourage
33 new innovative solutions (Kivimaa et al. 2017b) (See 7.4.8).

35 **7.4. Policy Instruments for Land and Climate**

36 This section outlines policy responses to risk. It describes multi-level policy instruments (7.4.1),
37 policy instruments for social protection (7.4.2), policies responding to hazard (7.4.3), GHG fluxes
38 (7.4.4), desertification (7.4.5), land degradation (7.4.6), economic instruments (7.4.7), enabling
39 effective policy instruments through policy mixes (7.4.8), and barriers to sustainable land
40 management and overcoming these barriers (7.4.9).

41 Policy instruments are used to influence behaviour and affect a response to do, not do, or continue to
42 do certain things (Anderson 2010) and can be invoked at multiple levels (international, national,
43 regional, and local) by multiple actors (See Table 7.2). For efficiency, equity and effectiveness
44 considerations, the appropriate choice of instrument for the context is critical, and across the topics
45 addressed in this report the instruments will vary considerably. A key consideration is whether the

1 benefits of the action will generate private or public social net benefits. Pannell (2008) provides a
2 widely-used framework for identifying the appropriate type of instrument depending on whether the
3 actions encouraged by the instrument are private or public, and positive or negative. Positive
4 incentives (such as financial or regulatory instruments) are appropriate where the public net benefits
5 are highly positive and the private net benefits are close to zero. This is likely to be the case for GHG
6 mitigation measures such as carbon pricing. Many other GHG mitigation measures (more effective
7 water or fertiliser use, better agricultural practices, less food waste, agroforestry systems, better forest
8 management) discussed in previous chapters may have substantial private as well as public benefit.
9 Extension (knowledge provision) is recommended for when public net benefits are highly positive and
10 private net benefits slightly positive, again for some GHG mitigation measures, and many adaptations,
11 food security and sustainable land management measures. Where the private net benefits are slightly
12 positive but the public net benefits highly negative, negative incentives (such as regulations and
13 prohibitions) are appropriate, for example over-application of fertiliser.

14 While Pannell (2008)'s framework is useful, it does not address considerations relating to the time-
15 scale of actions and their consequences particularly in the long time-horizons involved under climate
16 change: private benefits may accrue in the short term but become negative over time (Outka 2012)
17 and some of the changes necessary will require transformation of existing systems (Park et al. 2012;
18 Hadarits et al. 2017) for which a more comprehensive suite of instruments would be necessary.
19 Furthermore, the framework applies to private land ownership, so where land is in different ownership
20 structures, different mechanisms will be required. Indeed, land tenure is recognised as a factor in
21 barriers to Sustainable Land Management and an important Governance consideration (see 7.4.9,
22 7.6.4). A thorough analysis of the implications of policy instruments temporally, spatially and across
23 other sectors and goals (e.g., climate v. development) is essential before implementation to avoid
24 unintended consequences and achieve policy coherence (7.4.8).

25

26 **7.4.1. Multi-level Policy Instruments**

27 Policy responses and planning in relation to land and climate interactions occur at and across multiple
28 levels, involve multiple actors, and utilise multiple planning mechanisms (Urwin and Jordan 2008).
29 Climate change is occurring on a global scale while the impacts of climate change vary from region to
30 region and even within a region. Therefore, in addressing local climate impacts, local governments
31 and communities are key players. Advancing governance of *climate change* across all levels of
32 government and relevant stakeholders is crucial to avoid policy gaps between local action *plans* and
33 national/sub-national policy frameworks (Corfee-Morlot et al. 2009).

34 This section of the chapter identifies policies by level that respond to land and climate problems and
35 risks. As risk management in relation to land and climate occurs at multiple levels by multiple actors,
36 and across multiple sectors in relation to hazards (as listed on Table 7.2), risk governance, or the
37 consideration of the landscapes of risk arising from Chapters 2 through 6 is addressed in Sections 7.5
38 and 7.6. Categories of instruments include regulatory instruments (command and control measures),
39 economic and market instruments (creating a market, sending price signals, or employing a market
40 strategy), voluntary or persuasive instruments (persuading people to internalise behaviour), and
41 managerial (arrangements including multiple actors in cooperatively administering a resource or
42 overseeing an issue) (Gupta et al. 2013a; Hurlbert 2018b).

43 Given the complex spatial and temporal dynamics of risk, a comprehensive, portfolio of instruments
44 and responses is required to comprehensively manage risk. Operationalising a portfolio response can
45 mean layering, sequencing or integrating approaches. Layering means that within a geographical area,
46 households are able to benefit from multiple interventions simultaneously (e.g., those for family
47 planning and those for livelihoods development). A sequencing approach starts with those

1 interventions, which address the initial binding constraints, and then further interventions are later
2 added (e.g., the poorest households first receive grant-based support before then gaining access to
3 appropriate microfinance or market-oriented initiatives). Integrated approaches involve cross-sectoral
4 support within the framework of one program (Scott et al. 2016; Tengberg and Valencia 2018) (see
5 7.4.8, 7.5.6, and 7.6.3).

6 Climate related risk could be categorised by climate impacts such as flood, drought, cyclone etc.
7 (Christenson et al. 2014). Table 7.2 outlines instruments relating to impacts responding to the risk of
8 climate change, food insecurity, land degradation and desertification, and hazards (flood, drought,
9 forest fire), and GHG fluxes (climate mitigation).

1

Table 7.2 Policies/Instruments that address multiple land-climate risks at different jurisdictional levels

| Scale | Policy/Instrument | Food Security | Land degradation & desertification | Sustainable land management | Climate related Extremes | GHG flux climate change mitigation |
|---------------------------------|--|---------------|------------------------------------|-----------------------------|--------------------------|------------------------------------|
| Global/ Cross Border | Finance mechanisms (also National) | X | X | X | X | X |
| | Certification (also National) | | X | X | | X |
| | Standards (including Risk Standards)(also National) | | X | X | X | X |
| | Market based systems (also National) | | | X | | X |
| | Payments for Ecosystem Services (also National) | | X | X | X | X |
| | Disaster assistance (also National) | | | | X | |
| National | Taxes | X | | X | | X |
| | Subsidies | X | X | X | | X |
| | Direct Income Payments (with Cross-Compliance) | X | X | X | | X |
| | Border adjustments (e.g., tariffs) | X | | | | X |
| | Grants | X | X | X | X | X |
| | Bonds | X | X | X | | X |
| | Forecast-based finance, targeted microfinance | X | X | X | | X |
| | Insurance (various forms) | X | | | X | |
| | Hazard information and communication (also sub-national and local) | X | | | X | |
| | Drought preparedness plans (also sub-national and local) | X | | | X | |
| | Fire policy (suppression or prescribed fire management) | | | X | X | X |
| | Regulations | X | X | X | X | X |
| | Land ownership laws (reform of, if necessary, for secure land title, or access/control) | X | X | X | | |
| | Protected Area Designation and management | | X | X | | |
| | Extension – including skill and community development for livelihood diversification (also sub-national and local) | X | X | X | X | X |
| Sub-national | Spatial and landuse planning | X | X | | X | |
| | Watershed management | X | X | | | |
| Local | Landuse zoning, spatial planning and integrated landuse planning | X | | X | X | |
| | Community-based awareness programmes | X | X | X | X | X |

2

This table highlights policy and instruments addressing key themes identified in this chapter;

3

an X indicates the relevance of the policy or instrument to the corresponding theme.

4

7.4.2. Policies for Food Security and Social Protection

There is *medium evidence* and *high agreement* that a combination of structural and non-structural policies is required in averting and minimising as well as responding to land and climate change risk, including food and livelihood security. If disruptions to elements of food security are long-lasting, policies are needed to change practices

If disruptions to food and livelihood systems are temporary, then policies aimed at stemming worsening human wellbeing and stabilising short-term income fluctuations in communities (such as increasing rural credit or providing social safety net programs) may be appropriate (Ward 2016).

7.4.2.1. Policies to ensure availability, access, utilisation, and stability of food

Food security is affected by interactions between climatic factors (rising temperatures, changes in weather variability and extremes), changes in land-use and land degradation, and socio-economic pathways and policy choices related to food systems (see Figure 7.1 and Figure 7.2). As outlined in Chapter 5, key aspects of food security are food availability, access to food, utilisation of food, and stability of food systems.

While comprehensive reviews of policy are rare and additional data is needed (Adu et al. 2018), evidence indicates the result of food security interventions vary widely due to differing values underlying the design of instruments. A large portfolio of measures is available to shape outcomes in these areas from the use of tariffs or subsidies to payments for production practices (OECD 2018). In the past, efforts to increase food production through significant investment in agricultural research including crop improvement have benefited farmers by increasing yields and reducing losses, and have helped consumers by lowering food prices (Pingali 2012, 2015; Alston and Pardey 2014; Popp et al. 2013). Public spending on agriculture research and development has been more effective at raising sustainable agriculture productivity than irrigation or fertiliser subsidies (OECD 2018). Yet, on average between 2015 and 2017, governments spent only around 14% of total agricultural support on services which includes physical and knowledge infrastructure, transport and ICT.

In terms of increasing food availability and supply, producer support, including policies mandating subsidies or payments, have been used to boost production of certain commodities or protect ecosystem services. Incentives can distort markets and farm business decisions in both negative and positive ways. For example, the European Union promotes meat and dairy production through voluntary coupled direct payments. These do not yet internalise external damage to climate, health, and groundwater (Velthof et al. 2014; Bryngelsson et al. 2016). In most countries, producer support has been declining since the mid-1990s (OECD 2018). Yet new evidence indicates that a government policy supporting producer subsidy could encourage farmers to adopt new technologies and reduce GHG reductions in agriculture (*medium evidence, high agreement*). However, this will require large capital (Henderson 2018). Since a 1995 reform in its Forest Law, Costa Rica has effectively used a combination of fuel tax, water tax, loans and agreements with companies, to pay landowners for agroforestry, reforestation and sustainable forest management (Porrás and Asquith 2018).

Inland capture fisheries and aquaculture are an integral part of nutrition security and livelihoods for large numbers of people globally (Welcomme et al. 2010; Hall et al. 2013; Tidwell and Allan 2001; Youn et al. 2014) and are increasingly vulnerable to climate change and competing land and water use (Allison et al. 2009; Youn et al. 2014). Future production may increase in some high-latitude regions (*low confidence*) but production is likely to decline in low latitude regions under future warming (*high confidence*) (Brander and Keith 2015; Brander 2007). However over-exploitation and degradation of rivers has resulted in a decreasing trend in contribution of capture fisheries to protein security in comparison to managed aquaculture (Welcomme et al. 2010). Aquaculture however competes for land and water resources with many negative ecological and environmental impacts (Verdegem and Bosma 2009; Tidwell and Allan 2001). Inland capture fisheries are undervalued in national and regional food security, ecosystem services and economy, are data deficient and are neglected in terms of supportive policies at national levels and absent in Sustainable Development Goals (Cooke et al. 2016; Hall et al. 2013; Lynch et al. 2016). Revival of sustainable capture fisheries and converting aquaculture to

1 environmentally less damaging management regimes is likely to succeed by investment in
2 recognition of their importance, improved valuation and assessment, secure tenure and adoption of
3 social, ecological and technological guidelines besides upstream-downstream river basin cooperation
4 and maintenance of ecological flow regimes in rivers (Youn et al. 2014; Mostert et al. 2007; Ziv et al.
5 2012; Hurlbert and Gupta 2016; Poff et al. 2003; Thomas 1996; FAO 2015a).

6 Extension services, and policies supporting agricultural extension systems, are also critical.
7 Smallholder farmer-dominated agriculture is currently the backbone of global food security in the
8 developing world. Without education and incentives to manage land and forest resources in a manner
9 that allows regeneration of both the soils and wood stocks, smallholder farmers tend to generate
10 income through inappropriate land management practices, engage in agricultural production on
11 unsuitable land and use fertile soils, timber and firewood for brick production and construction and
12 secondly engage in charcoal production (deforestation) as a coping mechanism (increasing income)
13 against food deficiency (Munthali and Murayama 2013). Through extension services, governments
14 can play a proactive role in providing information on climate and market risks, animal and plant
15 health. Farmers with greater access to extension training retain more crop residues for mulch on their
16 fields (Jaleta et al. 2015, 2013; Baudron et al. 2014).

17 Food security cannot be achieved by increasing food availability alone. Policy instruments, which
18 increase access to food at the household level, include safety net programming and universal basic
19 income. The graduation approach, developed and tested over the past decade using randomised
20 control trials in six countries, has lasting positive impacts on income, as well as food and nutrition
21 security (Banerjee et al. 2015; Raza and Poel 2016) (*robust evidence, high agreement*). The
22 graduation approach layers and integrates a series of interventions designed to help the poorest:
23 consumption support in the form of cash or food assistance, transfer of an income generating asset
24 (such as a livestock) and training on how to maintain the asset, assistance with savings and coaching
25 or mentoring over a period of time to reinforce learning and provide support. Due to its success, the
26 graduation approach is now being scaled up, now used in over 38 countries and included by an
27 increasing number of governments in social safety-net programs (Hashemi, S.M. and de Montesquiou
28 2011).

29 At the national and global level, food price and trade policies impact access to food. Fiscal policies,
30 such as taxation, subsidies, or tariffs, can be used to regulate production and consumption of certain
31 foods and can affect environmental outcomes. In Denmark, tax on saturated fat content of food
32 adopted to encourage healthy eating habits accounted for 0.14% of total tax revenues between 2011
33 and 2012 (Sassi et al. 2018). A global tax on GHG emissions for example has large mitigation
34 potential and will generate tax revenues, but may also result in large reductions in agricultural
35 production (Henderson 2018). Consumer-level taxes on GHG intensive food may be applied to
36 address competitiveness issues between different countries, if some countries use taxes while others
37 do not. However, increases in prices might impose disproportionate financial burdens on low-income
38 households, and may not be publicly acceptable. A study examining the relationship between food
39 prices and social unrest found that between 1990 and 2011, food price increases have led to increases
40 in social unrest, whereas food price volatility has not been associated with increases in social unrest
41 (Bellemare 2015).

42 Interventions that allow people to maximise their productive potential while protecting the ecosystem
43 services may not ensure food security in all contexts. Some household land holdings are so small that
44 self-sufficiency is not possible (Venton 2018). Value chain development has in the past increased
45 farm income but delivered fewer benefits to vulnerable consumers (Bodnár et al. 2011). Ultimately, a
46 mix of production activities and consumption support is needed. Consumption support can be used to
47 help achieve the second important element of food security – access to food.

48 Agricultural technology transfer can help optimise food and nutrition security (see 7.4.4.3). Policies
49 that affect agricultural innovation span sectors and include “macro-economic policy-settings;
50 institutional governance; environmental standards; investment, land, labor and education policies; and
51 incentives for investment, such as a predictable regulatory environment and robust intellectual
52 property rights”.

1 The scientific community can partner across sectors and industries for better data sharing, integration,
2 and improved modelling and analytical capacities (Janetos et al. 2017; Lunt et al. 2016). To better
3 predict, respond to and prepare for concurrent agricultural failures, and gain a more systematic
4 assessment of exposure to agricultural climate risk, large data gaps need to be filled, as well as gaps in
5 empirical foundation and analytical capabilities (Janetos et al. 2017; Lunt et al. 2016). Data required
6 include global historical datasets, many of which are unreliable, inaccessible, or not available
7 (Maynard 2015; Lunt et al. 2016). Participation in co-design for scenario planning can build social
8 and human capital while improving understanding of food system risks and creating innovative ways
9 for collectively planning for a more equitable and resilient food system (Himanen et al. 2016; Meijer
10 et al. 2015; Van Rijn et al. 2012).

11 Demand management for food, including promoting healthy diets, reducing food loss and waste, is
12 covered in Chapter 5. There is a gap in knowledge regarding what policies and instruments support
13 demand management. There is *robust evidence and robust agreement* that changes in household
14 wealth and parents' education can drive changes in diet and improvements in nutrition (Headey et al.
15 2017). Bangladesh has managed to sustain a rapid reduction in the rate of child undernutrition for at
16 least two decades. Rapid wealth accumulation and large gains in parental education are the two largest
17 drivers of change (Headey et al. 2017). Educating consumers, and providing affordable alternatives,
18 will be critical to changing unsustainable food use habits relevant to climate change.

19 **7.4.2.2. Policies to secure social protection**

20 There is *medium evidence and high agreement* from all regions of the world that safety nets and social
21 protection schemes can provide stability which prevents and reduces abject poverty (Barrientos 2011;
22 Hossain 2018) (Cook and Pincus 2015; Huang and Yang 2017; Slater 2011; Sparrow et al. 2013;
23 Rodriguez-Takeuchi and Imai 2013; Bamberg et al. 2018) in the face of climatic stressors and land
24 change (Davies et al. 2013; Cutter et al. 2012b; Pelling 2011; Ensor 2011).

25 The World Bank estimates that globally social safety net transfers have reduced the absolute poverty
26 gap by 45% and the relative poverty gap by 16% (World Bank 2018). Adaptive social protection
27 builds household capacity to deal with shocks as well as the capacity of social safety nets to respond
28 to shocks. For low-income communities reliant on land and climate for their livelihoods and
29 wellbeing, social protection provides a way for vulnerable groups to manage weather and climatic
30 variability and deteriorating land conditions to household income and assets (*robust evidence, high
31 agreement*)(Baulch et al. 2006; Barrientos 2011; Harris 2013; Fiszbein et al. 2014; Kiendrebeogo et
32 al. 2017; Kabeer et al. 2010; FAO 2015b; Warner et al. 2018)(World Bank 2018).

33 Life cycle approaches to social protection are one approach, which some countries (such as
34 Bangladesh) are using when developing national social protection policies. These policies
35 acknowledge that households face risks across the life cycle from which they need to be protected. If
36 shocks are persistent, or occur numerous times, then policies can address concerns of a more
37 structural nature (Glauben et al. 2012). Barrett (2005), for example, distinguishes between the role of
38 safety nets (which include programs such as emergency feeding programs, crop or unemployment
39 insurance, disaster assistance, etc.) and cargo nets (which include land reforms, targeted microfinance,
40 targeted school feeding program, etc.). While the former prevents non-poor and transient poor from
41 becoming chronically poor, the latter is meant to lift people out of poverty by changing societal or
42 institutional structures. The graduation approach has adopted such systematic thinking with successful
43 results (Banerjee et al. 2015).

44 Social protection systems can provide buffers against shocks through vertical or horizontal expansion,
45 piggybacking on pre-established programmes, aligning social protection and humanitarian systems or
46 refocusing existing resources (Wilkinson et al. 2018; O'Brien, C.O., Scott, Z., Smith, G., Barca, V.,
47 Kardan, A., Holmes, R. Watson 2018); (Jones and Presler-Marshall 2015). There is increasing
48 evidence that forecast-based financing, linked to a social protection, can be used to enable
49 anticipatory actions based on forecast triggers and guaranteed funding ahead of a shock (Jjemba et al.
50 2018). Accordingly scaling up social protection based on an early warning could enhance timeliness,
51 predictability and adequacy of social protection benefits (Kuriakose et al. 2012; Costella et al. 2017a;

1 Wilkinson et al. 2018; O'Brien, C.O., Scott, Z., Smith, G., Barca, V., Kardan, A., Holmes, R. Watson
2 2018).

3 Countries at high-risk of natural disasters often have lower safety net coverage percent (World Bank
4 2018), and there is *medium evidence and medium agreement* that those countries with few financial
5 and other buffers have lower economic and social performance (Cutter et al. 2012b; Outreville
6 2011a). Social protection systems have also been seen as an unaffordable commitment of public
7 budget in many developing and low-income countries (Harris 2013). National systems may be
8 disjointed and piecemeal, and subject to cultural acceptance and competing political ideologies (Niño-
9 Zarazúa et al. 2012). For example, Liberia and Madagascar each have five different public works
10 programs, each with different donor organisations and different implementing agencies (Monchuk
11 2014). These implementation shortcomings mean that positive effects of social protection systems
12 might not be robust enough to shield recipients completely against the impacts of severe shocks or
13 from long-term losses and damages from climate change (*limited evidence, high agreement*) (Davies et
14 al. 2009; Umukoro 2013; Béné et al. 2012; Ellis et al. 2009).

15 There is increasing support for establishment of public-private safety nets to address climate related
16 shocks which are augmented by proactive preventative (adaptation) measures and related risk transfer
17 instruments that are affordable to the poor (Kousky et al. 2018b). Studies suggest that adaptive
18 capacity of communities have improved with regard to climate variability like drought when ex-ante
19 tools including insurance have been employed holistically; providing insurance in combination with
20 early warning and institutional and policy approaches that aim to reduce livelihood and food
21 insecurity as well as strengthen social structures (Shiferaw et al. 2014; Lotze-Campen and Popp 2012).
22 Bundling insurance with early warning and seasonal forecasting can reduce the cost of insurance
23 premiums (Daron and Stainforth 2014). The regional risk insurance scheme Africa Risk Capacity has
24 the potential to significantly reduce the cost of insurance premiums (Siebert 2016) while bolstering
25 contingency planning against food insecurity.

26 Work-for-insurance programs applied in the context of social protection have been shown to improve
27 livelihood and food security in Ethiopia (Berhane 2014; Mohammed et al. 2018) and Pakistan . The R4
28 Rural Resilience Program in Ethiopia is a widely cited example of a program that serves the most
29 vulnerable and includes aspects of resource management, and access by the poor to financial services
30 including insurance and savings (Linnerooth-bayer et al. 2018). Weather index insurance (such as
31 index based crop insurance) is being presented to low-income farmers and pastoralists in developing
32 countries (e.g., Ethiopia, India, Kazakhstan, China, South Asia) to complement informal risk sharing,
33 reducing the risk of lost revenue associated with variations in crop yield, and provide an alternative to
34 classic insurance (Bogale 2015a; Conradt et al. 2015; Dercon et al. 2014; Greatrex et al. 2015;
35 McIntosh et al. 2013). The ability of insurance to contribute to adaptive capacity depends on the
36 overall risk management and livelihood context of households — studies find that rain fed
37 agriculturalists and foresters with more years of education and credit but limited off-farm income are
38 more willing to pay for insurance than households who have access to remittances (such as from
39 family members who have migrated) (Bogale 2015a; Gan et al. 2014; Hewitt et al. 2017; Nischalke
40 2015). In Europe, modelling suggests that insurance incentives such as vouchers would be less
41 expensive than total incentivised damage reduction and may reduce residential flood risk by 12% in
42 Germany and 24% by 2040 (Hudson et al. 2016).

43 **7.4.3. Policies Responding to Climate Related Extremes**

44 **7.4.3.1. Risk Management Instruments**

45 Risk management addressing climate change has broadened to include mitigation, adaptation and
46 disaster preparedness in a process of risk management through instruments facilitating contingency
47 and cross sectoral planning (Hurlimann and March 2012; Oels 2013), social community planning, and
48 strategic, long term planning (Serrao-Neumann et al. 2015a). A comprehensive consideration
49 integrates principles from informal support mechanisms to enhance formal social protection
50 programming (Mobarak and Rosenzweig 2013; Stavropoulou et al. 2017) such that the social safety
51 net, disaster risk management, and climate change adaptation are all considered to enhance
52

1 livelihoods of the chronic poor (see char dwellers and recurrent floods in Jamuna and Brahmaputra
2 basins of Bangladesh (Awal 2013) (see also 7.4.7). Iterative risk management is an on-going process
3 of assessment, action, reassessment and response (Mochizuki et al. 2015) (see 7.5.2 and 7.4.7.2).

4 Important elements of risk planning include education, creation of hazard and risk maps; important
5 elements of predicting include hydrological and meteorological monitoring to forecast weather,
6 seasonal climate forecasts, aridity, flood and extreme weather; effective responding requires robust
7 communication systems that pass on information to enable response (Cools et al. 2016).

8 Gauging effectiveness of policy instruments is challenging. Timescale may influence outcomes. To
9 evaluate effectiveness researchers, program managers and communities strive to develop consistency,
10 comparability, comprehensiveness and coherence in their tracking. In other words, practitioners utilise
11 a consistent and operational conceptualisation of adaptation; focus on comparable units of analysis;
12 develop comprehensive datasets on adaptation action; and be coherent with our understanding of what
13 constitutes real adaptation (Ford and Berrang-Ford 2016). Increasing the use of systematic reviews or
14 randomised evaluations may also be helpful (Alverson and Zommers 2018).

15 Many risk management policy instruments are referred to by the International Organization of
16 Standardization which lists risk management principles, guidelines, and frameworks for explaining
17 the elements of an effective risk management program (ISO 2009). The standard provides practical
18 risk management instruments and makes a business case for risk management investments (McClellan
19 et al. 2010). Insurance addresses impacts associated with extreme weather events (storms, floods,
20 droughts, temperature extremes), but it can provide disincentives for reducing disaster risk at the local
21 level through the transfer of risk spatially to other places or temporally to the future (Cutter et al.
22 2012b) and uptake is unequally distributed across regions and hazards (Lal et al. 2012). Insurance
23 instruments (see 7.4.2 and 7.4.6) can take many forms (traditional indemnity based, market based crop
24 insurance, property insurance), and some are linked to livelihoods sensitive to weather as well as food
25 security (linked to social safety net programs) and ecosystems (coral reefs and mangroves). Insurance
26 instruments can also provide a framework for risk signals to adaptation planning and implementation
27 and facilitate financial buffering when climate impacts exceed current capabilities delivered through
28 both public and private finance (Bogale 2015b; Greatrex et al. 2015; Surminski et al. 2016). A
29 holistic consideration of all instruments responding to extreme impacts of climate change (drought,
30 flood etc.) is required when assessing if policy instruments are promoting livelihood capitals and
31 contributing to the resilience of people and communities (Hurlbert 2018b). This holistic consideration
32 of policy instruments leads to a consideration of risk governance (see 7.6).

33 Early warning systems are critical policy instruments for protecting lives and property, adapting to
34 climate change, and effecting adaptive climate risk management (*high confidence*) (Selvaraju 2011;
35 Cools et al. 2016; Travis 2013; Henriksen et al. 2018; Seng 2013; Kanta Kafle 2017; Garcia and
36 Fearnley 2012). Early warning systems exist at different levels and for different purposes including
37 the FAO global Information and Early Warning System (GIEWS) on food and agriculture, USAID
38 Famine, national and local extreme weather, species extinction, community based flood and landslide,
39 and informal pastoral drought early warning systems (Kanta Kafle 2017). Medium term warning
40 systems can identify areas of concern, hotspots of vulnerabilities and sensitivities, or critical zones of
41 land degradation (areas of concern)(see chapter 6) critical to reduce risks over five to ten years
42 (Selvaraju 2012). Early warning systems for dangerous climate shifts are emerging with
43 considerations of rate of onset, intensity, spatial distribution and predictability. Growing research in
44 the area is considering positive and negative lessons learned from existing hazard early warning
45 systems including lead time and warning response (Travis 2013).

46 For effectiveness, communication methods are best adapted to local circumstances, religious and
47 cultural based structures and norms, information technology, and local institutional capacity (Cools et
48 al. 2016; Seng 2013). Considerations of governance or the actors and architecture within the socio-

1 ecological system, is an important feature of successful early warning system development (Seng
2 2013). Effective early warning systems consider the critical links between hazard monitoring, risk
3 assessment, forecasting tools, warning and dissemination (Garcia and Fearnley 2012). These effective
4 systems incorporate local context by defining accountability, responsibility, acknowledging the
5 importance of risk perceptions and trust for an effective response to warnings. Although increasing
6 levels and standardisation nationally and globally is important, revising these systems through
7 participatory approaches cognizant of the tension with technocratic approaches improves success
8 (Cools et al. 2016; Henriksen et al. 2018; Garcia and Fearnley 2012).

9 **7.4.3.2. Drought related risk minimising instruments**

10 A more detailed review of drought instruments, and three broad policy approaches for responding to
11 drought, is provided in Cross-chapter Box 5: Case study on policy drought in Chapter 3. Three broad
12 approaches include: (1) early warning systems and response to the disaster of drought (through
13 instruments such as disaster assistance or crop insurance); (2) disaster response ex-ante preparation
14 (through drought preparedness plans); and (3) drought risk mitigation (proactive policies to improve
15 water use efficiency, make adjustments to water allocation, funds or loans to build technology such as
16 dugouts or improved soil management practices).

17 Drought plans are still predominantly reactive crisis management plans rather than proactive risk
18 management and reduction plans. Reactive crisis management plans treat only the symptoms and are
19 inefficient drought management practices. More efficient drought preparedness instruments are those
20 that address the underlying vulnerability associated with the impacts of drought thereby building
21 agricultural producer adaptive capacity and resilience (*high confidence*)(Cross-chapter Box 5: Case
22 study on policy drought, chapter 3).

23 **7.4.3.3. Fire related risk minimising instruments**

24 There is *robust evidence and high agreement* that fire strategies need to be tailored to site specific
25 conditions in an adaptive application that is assessed and reassessed over time (Dellasala et al. 2004;
26 Rocca et al. 2014). Strategies for fire management include fire suppression, prescribed fire and
27 mechanical treatments (such as thinning the canopy), and allowing wildfire with little or no active
28 management (Rocca et al. 2014). Fire suppression can degrade the effectiveness of forest fire
29 management in the long run (Collins et al. 2013).

30 Different forest types have different fire regimes and require different fire management policies
31 (Dellasala et al. 2004). For instance, Cerrado, a fire dependent savannah, utilises a fire management
32 policy different than the fire suppression policy (Durigan and Ratter 2016). The choice of strategy
33 depends on local considerations including land ownership patterns, dynamics of local meteorology,
34 budgets, logistics, federal and local policies, tolerance for risk and landscape contexts. In addition
35 there are trade-offs among the management alternatives and often no single management strategy will
36 simultaneously optimise ecosystem services including water quality and quantity, carbon
37 sequestration, or run off erosion prevention (Rocca et al. 2014).

38 **7.4.3.4. Flood related risk minimising instruments**

39 Flood risk management consists of command and control measures including spatial planning and
40 engineered flood defences (Filatova 2014), financial incentive instruments issued by regional or
41 national governments to facilitate cooperative approaches through local planning, enhancing
42 community understanding and political support for safe development patterns and building standards,
43 and regulations requiring local government participation and support for local flood planning (Burby
44 and May 2009). However, Filatova (2014) found that if autonomous adaptation is downplayed, people
45 are more likely to make land use choices that collectively lead to increased flood risks and leave costs
46 to governments. Taxes and subsidies that do not encourage (and even counter) perverse behaviour
47 (such as rebuilding in flood zones) are important instruments mitigating this cost to government.
48 Flood insurance has been found to be maladaptive as it encourages rebuilding in flood zones (O'Hare

1 et al. 2016)) and government flood disaster assistance negatively impacts average insurance coverage
2 the following year (Kousky et al. 2018a). Modifications to flood insurance can counter perverse
3 behaviour. One example is the provision of discounts on flood insurance for localities that undertake
4 one of 18 flood mitigation activities including structural mitigation (constructing dykes, dames, flood
5 control reservoirs), and non-structural initiatives such as point source control and watershed
6 management efforts, education and maintenance of flood-related databases (Zahran et al. 2010). Flood
7 insurance that provides incentives for flood mitigation, marketable permits and transferable
8 development rights (see case study of Flood and Food Security in Section 7.6) instruments can
9 provide price signals to stimulate autonomous adaptation, countering barriers of path dependency, and
10 the time lag between private investment decisions and consequences (Filatova 2014). To build
11 adaptive capacity, consideration needs to be made of policy instruments responding to flood including
12 flood zone mapping, land use planning, flood zone building restrictions, business and crop insurance,
13 disaster assistance payments, preventative instruments including environmental farm planning
14 (including soil and water management (see Chapter 6)) and farm infrastructure projects, and recovery
15 from debilitating flood losses ultimately through bankruptcy (Hurlbert 2018a). Non-structural
16 measures have been found to advance sustainable development as they are more reversible,
17 commonly acceptable and environmentally friendly (Kundzewicz 2002).

19 **7.4.4. Policies Responding to GHG fluxes**

20 **7.4.4.1. GHG fluxes and climate change mitigation**

21 Pathways reflecting current nationally stated mitigation ambitions as submitted under the Paris
22 Agreement would not limit global warming to 1.5°C with no or limited overshoot, but instead result
23 in a global warming of about 3°C by 2100 with warming continuing afterwards (IPCC 2018d).
24 Reversing warming after an overshoot of .2°C or larger during this century would require deployment
25 of CDR at rates and volumes that might not be achievable given considerable implementation
26 challenges (IPCC 2018d). This significant gap (Höhne et al. 2017; Rogelj et al. 2016) creates a
27 significant risk of global warming impacting land degradation, desertification, and food security (see
28 7.2;(IPCC 2018d). Action can be taken by 2030 adopting already known cost effective technology
29 (United Nations Environment Programme 2017), improving the finance, capacity building, and
30 technology transfer mechanisms of the UNFCCC, improving food security (listed by 73 nations in
31 their NDCs) and nutritional security (listed by 25 nations) (Richards, M., Bruun, T.B., Campbell,
32 B.M., Gregersen, L.E., Huyer 2015). UNFCCC Decision 1.CP21 reaffirmed the UNFCCC target that
33 ‘developed country parties provide USD 100 billion annually by 2020 for climate action in
34 developing countries’ (Rajamani 2011) and a new collective quantified goal above this floor is to be
35 set taking into account the needs and priorities of developing countries (Fridahl and Linnér 2016).

36 Mitigation policy instruments to address this shortfall include financing mechanisms, carbon pricing,
37 cap and trade or emissions trading, and technology transfer. While climate change is a global
38 commons problem containing free-riding problems, cost effective international policies that insure
39 countries get the most environmental benefit out of mitigation investments promote an international
40 climate policy regime (Nordhaus 1999; Aldy and Stavins 2012). Carbon pricing instruments may
41 provide an entry point for inclusion of agricultural appropriate carbon instruments. Models of cost
42 efficient distribution of mitigation across regions and sectors typically employ a global uniform
43 carbon price, but such treatment in the agricultural sector may impact food security (see 7.4.4.4).

44 One policy initiative to advance climate mitigation policy coherence (see 7.4.8) in this section is the
45 phase out of subsidies for fossil fuel production. The G20 agreed in 2009, and the G7 agreed in 2016,
46 to phase out these subsidies by 2025. Subsidies include lower tax rates or exemptions and rebates of

1 taxes on fuels used by particular consumers (diesel fuel used by farming, fishing etc.), types of fuel, or
2 how fuels are used. The OECD estimates the overall value of these subsidies to be between USD 160–
3 200 billion annually between 2010 and 2014 (OECD 2015). The phase out of fossil fuel subsidies has
4 important economic, environmental and social benefits. Coady et al. (2017) estimate the economic
5 and environmental benefits of reforming fossil fuel subsidies could be valued worldwide at USD 4.9
6 trillion in 2013, and USD 5.3 trillion in 2015. Eliminating subsidies could have reduced emissions by
7 21% and raised 4% of global GDP as revenue (in 2013) and improved social welfare (Coady et al.
8 2017).

9 Legal instruments addressing perceived deficiencies in climate change mitigation include human
10 rights and liability. Developments in attribution science are improving the ability to detect human
11 influence on extreme weather and Marjanac et al. (2017) argue this broadens the legal duty of
12 government, business and others to manage foreseeable harms and may lead to more climate change
13 litigation (Marjanac et al. 2017). Peel and Osofsky (2017) argue that courts are becoming increasingly
14 receptive to employ human rights claims in climate change lawsuits (Peel and Osofsky 2017); citizen
15 suits in domestic courts are not a universal phenomenon and even if unsuccessful, Estrin (2016)
16 concludes they are important in underlining the high level of public concern.

17 **7.4.4.2. Mitigation instruments**

18 Similar instruments for mitigation could be applied to the land sector as in other sectors, including
19 market-based measures such as taxes and cap and trade systems; as well as standards and regulations;
20 subsidies and tax credits; information instruments and management tools; R&D investment; and
21 voluntary compliance programmes, but few regions have implemented agricultural mitigation
22 instruments (Cooper et al. 2013). Existing regimes focus on subsidies, grants and incentives, and
23 voluntary offset programmes.

24 **Market-based instruments**

25 Although carbon pricing is recognised to be an important cost-effective instrument in a portfolio of
26 climate policies (Aldy et al. 2010) (*high evidence, high agreement*), as yet no country is exposing
27 their agricultural sector emissions to carbon pricing in any comprehensive way. A carbon tax, fuel
28 tax, and carbon markets (cap and trade system or Emissions Trading Scheme (ETS), or baseline and
29 credit schemes, and voluntary markets) are predominant policy instruments that implement carbon
30 pricing. The advantage of carbon pricing is environmental effectiveness at relatively low cost
31 (Baranzini et al. 2017; Fawcett et al. 2014) (*high evidence, high agreement*). Furthermore, carbon
32 pricing could be used to raise revenue to reinvest in public spending, either to help certain sectors
33 transition to lower carbon systems, or to invest in public spending unrelated to climate change. Both
34 of these options may make climate policies more attractive and enhance overall welfare (Siegmeier et
35 al. 2018), but there is as yet no evidence of the effectiveness of emissions pricing in agriculture
36 (Grosjean et al. 2018). There is however, a clear need for progress in this area as without effective
37 carbon pricing, the mitigation potential identified in chapters 5 and 6 of this report will not be realised
38 (Boyce 2018) (*high evidence, high agreement*).

39 The price may be set at the Social Cost of Carbon (the incremental impact of emitting an additional
40 tonne of CO₂, or the benefit of slightly reducing emissions), but estimates of the SCC vary widely and
41 are contested (Pezzey 2019) (*high evidence, high agreement*). An alternative to the SCC includes a
42 pathways approaches that sets an emissions target and estimates the Carbon prices required to achieve
43 this at the lowest possible cost (Pezzey 2019). Theoretically, higher costs throughout the entire
44 economy result in reduction of carbon intensity as consumers and producers adjust their decisions in
45 relation to prices corrected to reflect the climate externality (Baranzini et al. 2017).

46 Both carbon taxes and cap and trade systems can reduce emissions, but cap and trade systems are
47 generally more cost effective (*medium evidence, high agreement*) (Haites 2018a). In both cases, the

1 design of the system is critical to its effectiveness at reducing emissions (Bruvoll and Larsen 2004;
2 (Lin and Li 2011)) (*high evidence, high agreement*). The trading system allows the achievement of
3 emission reductions in the most cost-effective manner possible and results in a market and price on
4 emissions that create incentives for the reduction of carbon pollution. The way allowances are
5 allocated in a cap and trade system is critical to its effectiveness and equity. Free allocations can be
6 provided to trade-exposed sectors such as agriculture either through historic allocations or output
7 based; the choice of which has important implications (Quirion 2009). Output based allocations may
8 be most suitable for agriculture also minimising leakage risk (see below) (Grosjean et al. 2018)
9 (Quirion 2009). There is *medium evidence* and *high agreement* that properly designed, a cap and
10 trade system can be a powerful policy instrument (Wagner 2013) and may collect more rents than a
11 variable carbon tax (Siegmeier et al. 2018; Schmalensee and Stavins 2017).

12 In the land sector carbon markets are challenging to implement. Although several countries and
13 regions have ETSs in place (for example the EU, Switzerland, the Republic of Korea, Quebec in
14 Canada, California in the USA (Narassimhan et al. 2018)), none have included non-CO₂ (methane and
15 nitrous oxide) emissions from agriculture. New Zealand is the only country currently considering
16 ways to incorporate agriculture into its ETS (see Case Study on the New Zealand Emissions Trading
17 Scheme).

18 Three main reasons explain the lack of implementation to date:

19 1. The large number of heterogeneous buyers and sellers, combined with the difficulties of
20 monitoring, reporting and verification (MRV) of emissions from biological systems introduce
21 potentially high levels of complexity (and transaction costs). Effective policies therefore depend on
22 advanced MRV systems which are lacking in many (particularly developing) countries (Wilkes et al.
23 2017). This is discussed in more detail in the Case Study on the New Zealand Emissions Trading
24 Scheme.

25 2. Adverse distributional consequences (Grosjean et al. 2018) (*medium evidence, high agreement*).
26 Distributional issues depend, in part, on the extent that policy costs can be passed on to consumers,
27 and there is *medium evidence* and *medium agreement* that social equity can be increased through a
28 combination of non-market and market-based instruments (Haites 2018b).

29 3. Regulation, market-based or otherwise, adopted in only one jurisdiction and not elsewhere may
30 result in ‘leakage’ or reduced effectiveness – where production relocates to weaker regulated regions,
31 potentially reducing the overall environmental benefit. Although modelling studies indicate the
32 possibility of leakage following unilateral agricultural mitigation policy implementation (e.g.
33 Fellmann et al. 2018), there is no empirical evidence from the agricultural sector yet available.
34 Analysis from other sectors shows an overestimation of the extent of carbon leakage in modelling
35 studies conducted before policy implementation compared to evidence after the policy was
36 implemented (Branger and Quirion 2014). Options to avoid leakage include border adjustments
37 (emissions in non-regulated imports are taxed at the border, and payments made on products exported
38 to non-regulated countries are rebated); differential pricing for trade-exposed products and; output
39 based allocation (which effectively works as a subsidy for trade-exposed products). Modelling shows
40 that border adjustments are the most effective at reducing leakage, but may exacerbate regional
41 inequality (Böhringer et al. 2012) and through their trade-distorting nature may contravene WTO
42 rules. The opportunity for leakage would be significantly reduced ideally through multi-lateral
43 commitments (Fellmann et al. 2018) (*medium evidence, high agreement*) but could also be reduced
44 through regional or bi-lateral commitments within trade agreements.

45

46

Case study: Including agriculture in the Emissions Trading Scheme in New Zealand

New Zealand has a high proportion of agricultural emissions at 49% (Ministry of the Environment 2018) - the next highest developed country agricultural emitter is Ireland at around 32% (EPA 2018) - and is considering to incorporate agricultural non-CO₂ gases into the existing national ETS. In the original design of the ETS in 2008, agriculture was intended to be included from 2013, but successive Governments deferred the inclusion (Kerr and Sweet 2008) due to concerns about competitiveness, lack of mitigation options and the level of opposition from those potentially affected (Cooper and Rosin 2014). Now though, as the country's agricultural emissions are 12% above 1990 levels, and the country's total gross emissions have increased 19.6% above 1990 levels (New Zealand Ministry for the Environment 2018), there is a recognition that without any targeted policy for agriculture, only 52% of the country's emissions face any substantive incentive to mitigate (Narassimhan et al. 2018). Including agriculture in the ETS is one option to provide incentives for emissions reductions in that sector. Other options are discussed in Section 7.4.4. Although some producer groups raise concern that including agriculture will place New Zealand producers at a disadvantage compared with their international competitors who do not face similar mechanisms (New Zealand Productivity Commission 2018), there is generally greater acceptance of the need for climate policies for agriculture.

The inclusion of non-CO₂ emissions from agriculture within an ETS is potentially complex however, due to the large number of buyers and sellers if obligations are placed at farm level, and different choices of how to estimate emissions from biological systems in cost-effective ways. New Zealand is currently investigating practical and equitable approaches to include agriculture through advice being provided by the Interim Climate Change Committee (ICCC 2018). Main questions centre around the point of obligation for buying and selling credits, where trade-offs have to be made between providing incentives for behaviour change at farm level and the cost and complexity of administering the scheme (Agriculture Technical Advisory Group 2009; Kerr and Sweet 2008). The two potential points of obligation are at the processor level or at the individual farm level. Setting the point of obligation at the processor level means that farmers would face limited incentive to change their management practices, unless the processors themselves rewarded farmers for lowered emissions. Setting it at the individual farm level would provide a direct incentive for farmers to adopt mitigation practices, however the reality of having thousands of individual points of obligation would be administratively complex and could result in high transaction costs (Beca Ltd 2018).

Monitoring, reporting and verification (MRV) of agricultural emissions presents another challenge especially if emissions have to be estimated at farm level. Again, trade-offs have to be made between accuracy and detail of estimation method and the complexity, cost and audit of verification (Agriculture Technical Advisory Group 2009).

The ICCC is also exploring alternatives to an ETS to provide efficient abatement incentives (ICCC 2018).

Some discussion in New Zealand also focuses on a differential treatment of methane compared to nitrous oxide, Methane is a short-lived gas with a perturbation lifetime of twelve years in the atmosphere; nitrous oxide on the other hand is a long-lived gas and remains in the atmosphere for 114 years (Allen et al. 2016). Long-lived gases have a cumulative and essentially irreversible effect on the climate (IPCC 2014b) so their emissions need to reduce to net-zero in order to avoid climate change. Short-lived gases however could potentially be reduced to a certain level and then stabilised and would not contribute further to warming, leading to suggestions of treating these two gases separately in the ETS or alternative policy instruments, possibly setting different budgets and targets for each (New Zealand Productivity Commission 2018). Reisinger et al. (2013) demonstrate that different metrics can have important implications globally and potentially at national and regional scales on the costs and levels of abatement.

While the details are still being agreed on in New Zealand, almost 80% of NDCs committed to action on mitigation in agriculture (FAO 2016), so countries will be looking for successful examples.

1 Australia's Emissions Reduction Fund, and the preceding Carbon Farming Initiative, are an example
2 of a baseline-and-credit scheme, which set an emissions intensity baseline and creates credits for
3 activities that generate emissions below the baseline, effectively a subsidy (Freebairn 2016). It is a
4 voluntary scheme, and has potential to create real and additional emission reductions through projects
5 reducing emissions and sequestering carbon (Verschuuren 2017) (*low evidence, low agreement*). Key
6 success factors in the design of such an instrument are policy-certainty for at least ten to twenty years,
7 regulation that focuses on projects and not uniform rules, automated systems for all phases of the
8 projects, and a wider focus of the carbon farming initiative on adaptation, food security, sustainable
9 farm business, and creating jobs (Verschuuren 2017). A recent review highlighted the issue of
10 permanence and reversal, and recommended that projects detail how they will maintain carbon in their
11 projects and deal with the risk of fire.

12 **7.4.4.3. Technology transfer and land use sectors**

13 Technology transfer has been part of the UNFCCC process since its inception and is a key element of
14 international climate mitigation and adaptation efforts under the Paris Agreement. The IPCC
15 definition of Technology transfer includes transfer of knowledge and technological cooperation (see
16 Glossary) and can include modifications to suit local conditions and/or integration with indigenous
17 technologies (Metz et al. 2000). This definition suggests greater heterogeneity in the applications for
18 climate mitigation and adaptation, especially in land use sectors where indigenous knowledge may be
19 important for long-term climate resilience Nyong et al. (2007). For land use sectors, the typical
20 reliance on trade and patent data for empirical analyses is generally not feasible as the "technology" in
21 question is often related to resource management and is neither patentable nor tradable (Glachant and
22 Dechezleprêtre 2017) and ill-suited to provide socially beneficially innovation for poorer farmers in
23 developing countries (Lybbert and Sumner 2012; Baker, Dean; Jayadev, Arjun; Stiglitz 2017).

24 Technology transfer has contributed to emissions reductions (*medium confidence*). A detailed study
25 for nearly 4000 Clean Development Mechanism (CDM) projects showed that 39% of projects had a
26 stated and actual technology transfer component, accounting for 59% of emissions reductions;
27 however, the more land-intensive projects (e.g., afforestation, bioenergy) showed lower percentages
28 (Murphy et al. 2015). Bioenergy projects that rely on agricultural residues offer substantially more
29 development benefits than those based on industrial residues from forests (Lee and Lazarus 2013).
30 Energy projects tended to have a greater degree of technology transfer under the CDM compared to
31 non-energy projects (Gandenberger et al. 2016). However, longer-term cooperation and collaborative
32 R&D approaches to technology transfer will be more important in land use sectors (compared to
33 energy or industry) due to the time needed for improved resource management and interaction
34 between researchers, practitioners and policy-makers. These approaches offer longer-term technology
35 transfer that is more difficult to measure compared to specific cooperation projects; empirical research
36 on the effects of R&D collaboration could help to avoid the "one-policy-fits-all" approach (Ockwell
37 et al. 2015).

38 There is increasing recognition of the role of technology transfer in climate adaptation, but in the land
39 use sector there are inherent adoption challenges specific to adaptation, due to uncertainties arising
40 from changing climatic conditions, agricultural prices, and suitability under future conditions (Biagini
41 et al. 2014). Engaging the private sector is important, as adoption of new technologies can only be
42 replicated with significant private sector involvement (Biagini and Miller 2013).

43 **7.4.4.4. International Cooperation under the Paris Agreement**

44 New cooperative mechanisms under the Paris Agreement illustrate the shift away from the Kyoto
45 Protocol's emphasis on obligations of developed country Parties to pursue investments and
46 technology transfer, to a more pragmatic, decentralised and collaborative approach (Savaresi 2016;
47 Jiang et al. 2017). These approaches can effectively include any combination of measures or
48 instruments related to adaptation, mitigation, finance, technology transfer and capacity-building,

1 which could be of particular interest in land use sectors where such aspects are more intertwined than
2 in energy or industry sectors. Article 6 sets out several options for international cooperation (Gupta
3 and Dube 2018).

4 The close relationship between emission reductions, adaptive capacity, food security and other
5 sustainability and governance objectives in the land sectors means that Article 6 could bring co-
6 benefits that increase its attractiveness and the availability of finance, while also bringing risks that
7 need to be monitored and mitigated against, such as uncertainties in measurements and the risk of
8 non-permanence (Thamo and Pannell 2016; Olsson et al. 2016; Schwartz et al. 2017). There has been
9 progress in accounting for land-based emissions, mainly forestry and agriculture (*medium evidence*,
10 *low agreement*), but various challenges remain (Macintosh 2012; Pistorius et al. 2017; Krug 2018).

11 Like the Clean Development Mechanism (CDM) and other existing carbon trading mechanisms,
12 participation in Article 6.2 and 6.4 of the Paris Agreement requires certain institutional and data
13 management capacities in the land sector to effectively benefit from the cooperation opportunities
14 (Totin et al. 2018). While the rules for the implementation of the new mechanisms are still under
15 development, lessons from REDD+ may be useful, which is perceived as more democratic and
16 participative than the CDM (Maraseni and Cadman 2015). Experience with REDD+ programs
17 emphasise the necessity to invest in “readiness” programs that assist countries to engage in strategic
18 planning and build management and data collection systems to develop the capacity and infrastructure
19 to participate in REDD+ (Minang et al. 2014). The overwhelming majority of countries (93%) cite
20 weak forest sector governance and institutions in their applications for REDD+ readiness funding
21 (Kissinger et al. 2012). Technology transfer for advanced remote sensing technologies that help to
22 reduce uncertainty in monitoring forests helps to achieve REDD+ “readiness” (Goetz et al. 2015).

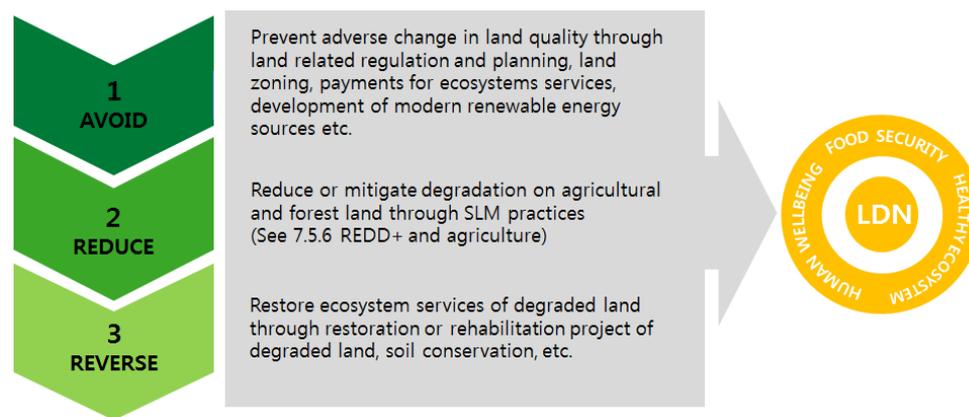
23 As well as new opportunities for finance and support, the Paris cooperation mechanisms and the
24 associated roles for technology transfer bring new challenges, particularly in reporting, verifying and
25 accounting in land use sectors. Since developing countries must now achieve, measure and
26 communicate emission reductions, they now have value for both developing and developed countries
27 in achieving their NDCs, but reductions cannot be double-counted (i.e., towards multiple NDCs). All
28 countries have to prepare and communicate NDCs, and many countries have included in their NDCs
29 either economy-wide targets that include the land use sectors, or specific targets for the land use
30 sectors. The Katowice climate package clarifies that all Parties have to submit ‘Biennial
31 Transparency Reports’ from 2024 onwards using common reporting formats, following most recent
32 IPCC Guidelines (use of the 2013 Supplement on Wetlands is encouraged), identifying key categories
33 of emissions, ensuring time-series consistency, and providing completeness and uncertainty
34 assessments as well as quality control (UNFCCC 2018a; Schneider and La Hoz Theuer 2019). In
35 total, the ambiguity in how countries incorporate land use sectors into their NDC is estimated to lead
36 to an uncertainty of more than 2 GtCO₂ in 2030 (Fyson and Jeffery 2018). Uncertainty is lower if the
37 analysis is limited to countries that have provided separate land use sector targets in their NDCs
38 (Benveniste et al. 2018).

39 **7.4.5. Policies Responding to Desertification and Degradation – Land** 40 **Degradation Neutrality (LDN)**

41 Land degradation neutrality (LDN) (SDG Target 15.3), evolved from the concept of Net Zero Land
42 Degradation, which was introduced by the UNCCD to promote sustainable land management (Kust et
43 al. 2017; Stavi and Lal 2015; Chasek et al. 2015). Neutrality here implies no net loss of the land-based
44 natural resource and ecosystem services relative to a baseline or a reference state (UNCCD 2015;
45 Kust et al. 2017; Easdale 2016; Cowie et al. 2018a; Stavi and Lal 2015; Grainger 2015; Chasek et al.
46 2015). Land degradation neutrality can be achieved by reducing the rate of land degradation (and
47 concomitant loss of ecosystem services) and increasing the rate of restoration and rehabilitation of

1 degraded or desertified land. Therefore, the rate of global land degradation is not to exceed that of
 2 land restoration in order to achieve land degradation neutrality goals (adopted as national platform for
 3 actions by > 100 countries)(Stavi and Lal 2015; Grainger 2015; Chasek et al. 2015; Cowie et al.
 4 2018a; Montanarella 2015). Achieving land degradation neutrality would decrease the environmental
 5 footprint of agriculture, while supporting food security and sustaining human wellbeing (UNCCD
 6 2015; Safriel 2017; Stavi and Lal 2015; Kust et al. 2017).

7 Response hierarchy - avoiding, reducing and reversing land degradation - is the main policy response
 8 (Chasek et al. 2019, Wonder and Bodle 2019, Cowie et al. 2018, Orr et al. 2017). The LDN response
 9 hierarchy encourages through regulation, planning and management instruments, the adoption of
 10 diverse measures to avoid, reduce and reverse land degradation in order to achieve LDN (Cowie et al.
 11 2018b; Orr et al. 2017).



12
 13 **Figure 7.4 LDN response hierarchy**

14 **Source: Adapted from (Liniger et al. 2019; UNCCD/Science-Policy-Interface 2016)**

15

16 Chapter 3 categorised policy responses into two categories; (1) avoiding, reducing and reversing it
 17 through sustainable land management; and (2) providing alternative livelihoods with economic
 18 diversification. Land degradation neutrality could be achieved through planned effective actions,
 19 particularly by motivated stakeholders those who play an essential role in a land-based climate change
 20 adaptation (Easdale 2016; Qasim et al. 2011; Cowie et al. 2018a; Salvati and Carlucci 2014). Human
 21 activities impacting the sustainability of drylands is a key consideration in adequately reversing
 22 degradation through restoration or rehabilitation of degraded land (Easdale 2016; Qasim et al. 2011;
 23 Cowie et al. 2018a; Salvati and Carlucci 2014).

24 LDN actions and activities play an essential role for a land-based approach to climate change
 25 adaptation (UNCCD 2015). Policies responding to degradation and desertification include improving
 26 market access, gender empowerment, expanding access to rural advisory services, strengthening land
 27 tenure security, payments for ecosystem services, decentralised natural resource management,
 28 investing into research and development, investing into monitoring of desertification and desert
 29 storms, developing modern renewable energy sources, investing into modern renewable energy
 30 sources, and developing and strengthening climate services. Policy supporting economic
 31 diversification include investing in irrigation, expanding agricultural commercialisation, and
 32 facilitating structural transformations in rural economies. (Chapter 3). Policies and actions also
 33 include promoting local and indigenous knowledge, soil conservation, agroforestry, crop-livestock
 34 interactions as an approach to manage land degradation, and forest based activities such as
 35 afforestation, reforestation, and changing forest management (Chapter 4). Measures identified for
 36 achievement of LDN include; effective financial mechanisms (for implementation of land restoration
 37 measures and the long-term monitoring of progress), parameters for assessing land degradation,

1 detailed plans with quantified objectives and timelines (Kust et al. 2017; Sietz et al. 2017; Cowie et al.
2 2018a; Montanarella 2015; Stavi and Lal 2015).

3 Implementing the international LDN target into national policies has been a challenge (Cowie et al.
4 2018a; Grainger 2015) as baseline land degradation or desertification information is not always
5 available (Grainger 2015) and challenges exist in monitoring LDN as it is a dynamic process (Sietz et
6 al. 2017; Grainger 2015; Cowie et al. 2018a). Wunder and Bodle (2019) propose that LDN be
7 implemented and monitored through indicators at the national level. Effective implementation of
8 global LDN will be supported by integrating lessons learned from existing programs designed for
9 other environmental objectives and closely coordinate LDN activities with actions for climate change
10 adaptation and mitigation at both global and national levels (*high confidence*) (Stavi and Lal 2015;
11 Grainger 2015).

12

13 **7.4.6. Policies Responding to Land Degradation**

14 **7.4.6.1. Land Use Zoning**

15 Land use zoning divides a territory (including local, sub-regional or national) into zones with different
16 rules and regulations for land use (mining, agriculture, urban development etc.), management
17 practices and land cover change (Metternicht 2018). While the policy instrument is zoning
18 ordinances, the process of determining these regulations is covered in integrated land use planning
19 (See 7.6.2). Urban zoning can guide new growth in urban communities outside current and
20 forecasted hazard areas, assist relocating existing dwellings to safer sites and manage postevent
21 redevelopment in ways to reduce future vulnerability (Berke and Stevens 2016). Holistic integration
22 of climate mitigation and adaptation are interdependent and can be implemented by restoring urban
23 forests, improving parks (Brown 2010; Berke and Stevens 2016). Zoning ordinances can contribute to
24 sustainable land management through protection of natural capital by preventing or limiting
25 vegetation clearing, avoiding degradation of planning for rehabilitation of degraded land or
26 contaminated sites, promoting conservation and enhancement of ecosystems and ecological corridors
27 (Metternicht 2018; Jepson and Haines 2014). Zoning ordinances can also encourage higher density
28 development, mixed use, local food production, encourage transportation alternatives (bike paths and
29 transit oriented development), preserve a sense of place, and increase housing diversity and
30 affordability (Jepson and Haines 2014). Conservation planning varies by context and may include one
31 or several adaptation approaches including protecting current patterns of biodiversity, large intact
32 natural landscapes, and geophysical settings. Conservation planning may also maintain and restore
33 ecological connectivity, identify and manage areas that provide future climate space for species
34 expected to be displaced by climate change, and identify and protect climate refugia (Stevanovic et
35 al. 2016; Schmitz et al. 2015).

36 Anguelovski et al. (2016) studied land use interventions in eight cities in the global north and south
37 and concluded that historic trends of socioeconomic vulnerability can be reinforced which could be
38 avoided with a consideration of the distribution of adaptation benefits and prioritising beneficial
39 outcomes for disadvantaged and vulnerable groups when making future adaptation plans.
40 Concentration of adaptation resources within wealthy business districts creating ecological enclaves
41 exacerbated climate risks elsewhere and building of climate adaptive infrastructure such as sea walls
42 or temporary flood barriers occurred at the expense of underserved neighbourhoods (Anguelovski et
43 al. 2016a).

44

45 **7.4.6.2. Conserving biodiversity and ecosystem services**

46 There is *limited evidence but high agreement* that ecosystem-based adaptation (biodiversity,
47 ecosystem services, and nature's contribution to people (see chapter 6)) and incentives for ecosystem

1 services (including PES) play a critical part of an overall strategy to help people adapt to the adverse
2 effects of climate change on land (UNEP 2009) (Bonan 2008; Millar et al. 2007; Thompson et al.
3 2009).

4
5 Ecosystem based adaptation can promote socio-ecological resilience by enabling people to adapt to
6 the impacts of climate change on land and reduce their vulnerability (Ojea 2015). Ecosystem based
7 adaptation can promote nature conservation while alleviating poverty and even provide co-benefits by
8 removing greenhouse gas (Scarano 2017) and protecting livelihoods (Munang et al. 2013). For
9 example, mangroves provide diverse ecosystem services such as carbon storage, fisheries, non-timber
10 forest products, erosion protection, water purification, shore-line stabilisation and also regulate storm
11 surge and flooding damages, thus enhancing resilience and reducing climate risk from extreme events
12 such as cyclones (Rahman, M.M., Khan, M.N.I., Hoque, A.K.F., Ahmed 2014; Donato et al. 2011;
13 Das and Vincent 2009; Ghosh et al. 2015; Ewel et al. 1998).

14
15 There has been considerable increase in the last decade of payments for ecosystem services (PES), or
16 programmes that exchange value for land management practices intended to ensure ecosystem
17 services (Salzman et al. 2018; Yang and Lu 2018; Barbier 2011). However, there is a deficiency in
18 comprehensive and reliable data concerning PES' impact on ecosystems, human well-being, their
19 efficiency, and effectiveness (Pynegar et al. 2018; Reed et al. 2014; Salzman et al. 2018; Barbier
20 2011; Yang and Lu 2018). While some studies assess ecological effectiveness and social equity,
21 fewer assess economic efficiency (Yang and Lu 2018). Part of the challenge surrounds the fact that
22 the majority of ecosystem services are not marketed, so determining how changes in ecosystems
23 structures, functions and processes influence the quantity and quality of ecosystem service flows to
24 people is challenging (Barbier 2011). PES include agri-environmental targeted outcome based
25 payments, but challenges exist in relation to scientific uncertainty, pricing, timing of payments,
26 increasing risk to land managers, World Trade Organization compliance, and barriers of land
27 management and scale (Reed et al. 2014).

28
29 PES is contested (Wang and Fu 2013; Czembrowski and Kronenberg 2016) (Perry 2015) for four
30 reasons: (1) understanding and resolving trade-offs between conflicting groups of stakeholders (Wam
31 et al. 2016) (Matthies et al. 2015); (2) knowledge and technology capacity (Menz et al. 2013); (3)
32 challenges integrating PES with economic and other policy instruments (Ring and Schröter-Schlaack
33 2011; Tallis et al. 2008)(Elmqvist et al. 2003; Albert et al. 2014); and (4) top down climate change
34 mitigation initiatives which are still largely carbon centric with limited opportunities for decentralised
35 ecological restoration at local and regional scales (Vijge and Gupta 2014).

36
37 These challenges and contestations can be resolved with the participation of people in establishing
38 PES thereby addressing trust issues, negative attitudes, and resolving trade-offs between issues (such
39 as retaining forests that consume water versus the provision of run off, or balancing payments to
40 providers versus cost to society) (Sorice et al. 2018; Matthies et al. 2015). Similarly, a 'co-
41 constructive' approach is used involving a diversity of stakeholders generating policy relevant
42 knowledge for sustainable management of biodiversity and ecosystem services at all relevant spatial
43 scales, by the current Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem
44 Services (IPBES) initiative (Díaz et al. 2015). Invasive species are also best identified and managed
45 with the participation of people through collective decisions, coordinated programs, and extensive
46 research and outreach to address their complex social-ecological impacts (Wittmann et al. 2016;
47 Epanchin-Niell et al. 2010).

48
49 Ecosystem restoration with co-benefits for diverse ecosystem services can be achieved through
50 passive restoration, passive restoration with protection and active restoration with planting (Birch et
51 al. 2010; Cantarello et al. 2010). Taking into account costs of restoration and co-benefits from bundles
52 of ecosystem services (carbon, tourism, timber), the benefit cost ratio of active restoration and passive
53 restoration with protection was always less than 1, suggesting that financial incentives would be
54 required. Passive restoration was the most cost-effective with BCR was generally between 1 and 100
55 for forest, grassland and shrubland restoration (TEEB 2009; Cantarello et al. 2010). Passive

1 restoration is generally more cost-effective but there is a danger that it could be confused with
2 abandoned land in the absence of secure tenure and long time period (Zahawi et al. 2014). Net Social
3 Benefits of degraded land restoration in dry regions range from about 200–700 USD per hectare
4 (Cantarello et al., 2010). Investments in active restoration could benefit from analyses of past land
5 use, the natural resilience of the ecosystem, and the specific objectives of each project (Meli et al.
6 2017). One successful example is the Working for Water initiative in South Africa that linked
7 restoration through removal of invasive species and enhancing water security (Milton et al. 2003).

8
9 Forest, water and energy cycle interactions and teleconnections such as contribution to rainfall
10 potentially (2.5.4) (Aragão 2012; Ellison et al. 2017; Paul et al. 2018; Spracklen et al. 2012) provide a
11 foundation for achieving forest-based adaptation and mitigation goals. They are however poorly
12 integrated in policy and decision making including PES.

13 14 15 **7.4.6.3. Standards and certification for sustainability of biomass and land use** 16 **sectors**

17 During the past two decades, standards and certification have emerged as important sustainability and
18 conservation instruments for agriculture, forestry, bioenergy, land use management and bio-based
19 products (Lambin et al. 2014; Englund and Berndes 2015; Milder et al. 2015; Giessen et al. 2016a;
20 Endres et al. 2015; Byerlee et al. 2015; van Dam et al. 2010). Standards are normally voluntary but
21 can also become obligatory through legislation. A standard provides specifications or guidelines to
22 ensure that materials, products, processes and services are fit for their purpose, whereas certification is
23 the procedure through which an accredited party confirms that a product, process or service is in
24 conformity with certain standards. Standards and certification are normally carried out by separate
25 organisations for legitimacy and accountability (see 7.6.6). The International Organization for
26 Standardization (ISO) is a key source for global environmental standards. Those with special
27 relevance for land and climate include a recent standard on combating land degradation and
28 desertification (ISO 2017) and an earlier standard on sustainable bioenergy and biomass use (ISO
29 2015; Walter et al. 2018). Both aim to support the long-term transition to a climate-resilient
30 bioeconomy; there is *medium evidence* on the sustainability implications of different bioeconomy
31 pathways, but *low agreement* as to which pathways are socially and environmentally desirable
32 (Priefer et al. 2017; Johnson 2017; Bennich et al. 2017a).

33 Table 7.3 provides a summary of selected standards and certification schemes with a focus on land use
34 and climate: the tickmark shows inclusion of different sustainability elements, with all recognising the
35 inherent linkages between the biophysical and social aspects of land use. Some certification schemes
36 and best practice guidelines are specific to a particular agriculture crop (e.g., soya, sugarcane) or a
37 tree (oil palm) while others are general. International organisations promote sustainable land and
38 biomass use through good practice guidelines, voluntary standards and jurisdictional approaches
39 (Scarlat and Dallemand 2011; Stattman et al. 2018a; ISEAL Alliance). Other frameworks, such as the
40 Global Bioenergy Partnership (GBEP) focus on monitoring land and biomass use through a set of
41 indicators that are applied across partner countries, thereby also promoting technology (knowledge)
42 transfer (GBEP 2017). The Economics of Land Degradation Initiative (ELD) provides common
43 guidelines for economic assessments of land degradation (Nkonya et al. 2013).

44 Whereas current standards and certification focus primarily on land, climate and biomass impacts
45 where they occur, more recent analysis considers trade-related land use change by tracing supply
46 chain impacts from producer to consumer, leading to the notion of “imported deforestation” that
47 occurs from increasing demand and trade in unsustainable forest and agriculture products, which is
48 estimated to account for 26% of all tropical deforestation (Pendrill et al. 2019). Research and
49 implementation efforts aim to improve supply chain transparency and promote commitments to “zero
50 deforestation” (Gardner et al. 2018a; Garrett et al. 2019; Newton et al. 2018; Godar and Gardner
51 2019; Godar et al. 2015, 2016). France has developed specific policies on imported deforestation that
52 are expected to eventually include a zero deforestation label (Government of France 2019).

Table 7.3 Selected standards and certification schemes and their components or coverage

| Acronym | Scheme, programme or standard | Commodity/process, relation to others | Type of mechanism | Environmental | | | | | | Socio-economic | | |
|-------------------|--|--|-----------------------|---------------|--------------|--------------|------|-----|-------|----------------------------------|-------------|----------------------------|
| | | | | GHG emissions | Biodiversity | Carbon stock | Soil | Air | Water | Land use management ^a | Land rights | Food security ^b |
| ISCC | International Sustainability & Carbon Certification | All feedstocks, all supply chains | Certification | √ | √ | √ | √ | √ | √ | √ | √ | √ |
| Bonsucro | BonsucroEU | Sugar cane and derived products | Certification | √ | √ | √ | √ | √ | √ | √ | √ | |
| RTRS | Roundtable on Responsible Soy EU | Soy based products | Certification | √ | √ | √ | √ | √ | √ | √ | √ | |
| RSB | Roundtable on Sustainable Biomaterials EU | Biomass for biofuels and biomaterials | Certification | √ | √ | √ | √ | √ | √ | √ | √ | √ |
| SAN | Sustainable Agriculture | Various agricultural crops and commodities; Linked to Rain Forest Alliance | Technical Network | | √ | √ | √ | √ | √ | √ | | |
| RSPO RED | Roundtable on Sustainable Palm Oil RED | Palm oil products | Certification | √ | √ | √ | √ | √ | √ | √ | √ | √ |
| PEFC | Programme for Endorsement of Forest Certification | Forest management | Certification | | √ | √ | √ | √ | √ | √ | √ | c |
| FSC | Forest Stewardship Council | Forest Management | Certification | | √ | √ | √ | √ | √ | √ | √ | |
| SBP | Sustainable Biomass Programme | woody biomass (e.g., wood pellets, wood chips); Linked to PEFC and FSC | Certification | √ | √ | √ | √ | √ | √ | √ | √ | |
| WOCAT | World Overview of Conservation Approaches and Technologies | Global network on sustainable land management | Best Practice Network | | | √ | √ | √ | √ | √ | | |
| ISO 13065: 2015 | Bioenergy | biomass and bioenergy, including conversion processes | Standard | √ | √ | √ | √ | √ | √ | √ | √ | √d |
| ISO 14055-1: 2017 | Land Degradation and Desertification | land use management, including restoration of degraded land | Standard | √ | | | | √ | √ | √ | √ | |

Source: Modified from (European Commission 2012; DIAZ-CHAVEZ 2015).

√ indicates that the issue is addressed in the standard or scheme

^a includes restoration of degraded land in some cases (especially ISO 14055-1)

^b where specifically indicated

^c reference to the RSB certification/standard

^d where specifically noted

1 The sustainability of biofuels and bioenergy has been in particular focus during the past decade or so
2 due to biofuel mandates and renewable energy policies in the U.S., EU and elsewhere (van Dam et al.
3 2010; Scarlat and Dallemand 2011). The European Union Renewable Energy Directive (EU-RED)
4 established sustainability criteria in relation to EU renewable energy targets in the transport sector
5 (European Commission 2012), which subsequently had impacts on land use and trade with third-party
6 countries (Johnson et al. 2012). In particular, the EU-RED marked a departure in the context of
7 Kyoto/UNFCCC guidelines by extending responsibility for emissions beyond the borders of final use,
8 and requiring developing countries wishing to sell into the EU market to meet the sustainability
9 criteria (Johnson 2011b). The recently revised EU-RED provides sustainability criteria that include
10 management of land and forestry as well as socio-economic aspects (European Union 2018; Faaij
11 2018; Stattman et al. 2018b). Standards and certification aim to address potential conflicts between
12 different uses of biomass and most schemes also consider co-benefits and synergies (see Cross-
13 chapter Box 7: Bioenergy and BECCS in mitigation scenarios, in Chapter 6). Bioenergy may offer
14 additional income and livelihoods to farmers as well as improvements in technical productivity and
15 multi-functional landscapes (Rosillo Callé and Johnson 2010a; Kline et al. 2017; Araujo Enciso et al.
16 2016). Results depend on the commodities involved, and also differ between rural and urban areas.

17 Analyses on the implementation of standards and certification for land and biomass use have focused
18 on their stringency, effectiveness and geographical scope as well as socio-economic impacts such as
19 land tenure, gender and land rights (Diaz-Chavez 2011; German and Schoneveld 2012; Meyer and
20 Priess 2014). The level of stringency and enforcement varies with local environmental conditions,
21 governance approaches and the nature of the feedstock produced (Endres et al. 2015; Lambin et al.
22 2014; Giessen et al. 2016b; Stattman et al. 2018b). There is *low evidence and low agreement* on how
23 the application and use of standards and certification has actually improved sustainability beyond the
24 local farm, factory or plantation level; the lack of harmonisation and consistency across countries that
25 has been observed, even within a common market or economic region such as the EU, presents a
26 barrier to wider market impacts (Endres et al. 2015; Stattman et al. 2018b; ISEAL Alliance). In the
27 forest sector, there is evidence that certification programmes such as FSC have reduced deforestation
28 in the aggregate as well as reducing air pollution (Miteva et al. 2015; Mcdermott et al. 2015).
29 Certification and standards cannot address global systemic concerns such as impacts on food prices or
30 other market-wide effects but rather are aimed primarily at insuring best practices in the local context.
31 More general approaches to certification such as the Gold Standard are designed to accelerate
32 progress toward the SDGs as well as the Paris Climate Agreement by certifying investment projects
33 while also emphasising support to governments (Gold Standard).

34

7.4.6.4. Energy access and biomass use

Access to modern energy services is a key component of SDG 7, with an estimated 1.1 billion persons lacking access to electricity while nearly three billion people relying on traditional biomass (fuelwood, agriculture residues, animal dung, charcoal) for household energy needs (IEA 2017). Lack of access to modern energy services is significant in the context of land-climate systems because heavy reliance on traditional biomass can contribute to land degradation, household air pollution and GHG emissions (see Cross Chapter box 12: Traditional Biomass use, in this Chapter). A variety of policy instruments and programmes have been aimed at improving energy access and thereby reducing the heavy reliance on traditional biomass (see Table 7.2); there is *high evidence and high agreement* that programmes and policies that reduce dependence on traditional biomass will have benefits for health and household productivity as well as reducing land degradation (see section 4.5.4) and GHG emissions (Bailis et al. 2015; Cutz et al. 2017a; Masera et al. 2015; Goldemberg et al. 2018a; Sola et al. 2016a; Rao and Pachauri 2017; Denton et al. 2014). There can be trade-offs across different options, especially between health and climate benefits since more efficient wood stoves might have only limited effect, whereas gaseous and liquid fuels (e.g., biogas, LPG, bioethanol) will have highly positive health benefits and climate benefits that vary depending on specific circumstances of the substitution (Cameron et al. 2016; Goldemberg et al. 2018b). Unlike traditional biomass, modern bioenergy offers high quality energy services, although for household cookstoves, even the cleanest options using wood may not perform as well in terms of health and/or climate benefits (Fuso Nerini et al. 2017; Goldemberg et al. 2018b).

Case Study: Forest conservation instruments: REDD+ in the Amazon and India

Over 50 countries have developed national REDD+ strategies, which have key conditions for addressing deforestation and forest degradation (improved monitoring capacities, understanding of drivers, increased stakeholder involvement, and provided a platform to secure indigenous and community land rights), however to achieve its original objectives and to be effective under current conditions, forest-based mitigation actions need to be incorporated in national development plans and official climate strategies, and mainstreamed across sectors and levels of government (Angelsen et al. 2018a).

The Amazon region can illustrate the complexity of the implementation of REDD+, in the most biodiverse place of the planet, with millions of inhabitants and hundreds of ethnic groups, under the jurisdiction of eight countries. While different experiences can be drawn at different spatial scales, at the regional-level, for example, Amazon Fund (van der Hoff et al. 2018), at the subnational level (Furtado 2018), and at the local level (Alvarez et al. 2016; Simonet et al. 2019), there is *medium evidence and high agreement* that REDD+ has stimulated sustainable land-use investments but also is competing with other land uses (e.g., agroindustry) and scarce international funding (both public and private) (Bastos Lima et al. 2017b; Angelsen et al. 2018b)

In the Amazon, at the local level, a critical issue has been the incorporation of indigenous people in the planning and distribution of benefits of REDD+ projects. While REDD+, in some cases, has enhanced participation of community members in the policy-planning process, fund management, and carbon baseline establishment increased project reliability and equity (West 2016), it is clear that, in this region, insecure and overlapping land rights, as well as unclear and contradictory institutional responsibilities, are probably the major problems for REDD+ implementation (Loaiza et al. 2017). Despite legal and rhetoric recognition of indigenous land rights, effective recognition is still lacking (Aguilar-Støen 2017). The key to the success of REDD+ in the Amazon, has been the application of both, incentives and disincentives on key safeguard indicators, including land security, participation, and well-being (Duchelle et al. 2017).

1 On the other hand, at the subnational level, REDD+ has been unable to shape land-use dynamics or
2 landscape governance, in areas suffering strong exogenous factors, such as extractive industries, and
3 in the absence of effective regional regulation for sustainable land use (Rodriguez-Ward et al. 2018;
4 Bastos Lima et al. 2017b). Moreover, projects with weak financial incentives, engage households with
5 high off-farm income, which already are better off than the poorest families (Loaiza et al. 2015).
6 Beyond, operational issues, clashing interpretations of results might bring clashes between
7 implementing countries or organisations and donor countries, which have revealed concerns over the
8 performance of projects (van der Hoff et al. 2018)

9 REDD+ Amazonian projects often face methodological issues, including how to assess the
10 opportunity cost among landholders, and informing REDD+ implementation (Kweka et al. 2016).
11 REDD+ based projects depend on consistent environmental monitoring methodologies for measuring,
12 reporting and verification and, in the Amazon, land cover estimates are crucial for environmental
13 monitoring efforts (Chávez Michaelsen et al. 2017).

14 In India forests and wildlife concerns are on the concurrent list of the Constitution since an
15 amendment in 1976 thus giving the central or federal government a strong role in matters related to
16 governance of forests. High rates of deforestation due to development projects led to the Forest
17 (Conservation) Act (1980) which requires central government approval for diversion of forest land in
18 any state or union territory.

19 Before 2006 forest diversion for development projects leading to deforestation needed the forest
20 clearance from the Central Government under the provisions of the Forest (Conservation Act) 1980.
21 In order to regulate forest diversion and as payment for ecosystem services a Net Present Value
22 (NPV) frame-work was introduced by the Supreme Court of India informed by the Kanchan Chopra
23 committee (Chopra 2017). The Supreme Court established the Compensatory Afforestation
24 Management and Planning Authority (CAMPA) under which the fund collected for compensatory
25 afforestation and on account of NPV from project developers is deposited. The Forest (Conservation)
26 Act of 1980 does require compensatory afforestation in lieu of forest diversion and in addition after
27 CAMPA the payment of NPV to get the forest clearance for diversion has been added.

28 As of February 2018, USD 6,825 million had accumulated in CAMPA funds in lieu of NPV paid by
29 developers diverting forest land throughout India for non-forest use. Funds are released by the central
30 government to state governments out of this fund for afforestation and conservation related activities
31 to “compensate” for diversion of forests. This is now governed by legislation called CAMPA Act
32 passed by the Parliament of India in July 2016. The CAMPA mechanism has however invited
33 criticism on various counts in terms of undervaluation of forest, inequality, lack of participation and
34 environmental justice (Temper and Martinez-Alier 2013).

35 The other significant development related to forest land was the landmark legislation called the
36 Scheduled Tribes and Other Traditional Forest Dwellers (Recognition of Forest Rights) Act, 2006 or
37 Forest Rights Act passed by the Parliament of India in 2007. This is the largest forest tenure legal
38 instrument in the world and attempted to undo a historical injustice to forest dwellers and forest
39 dependent communities whose traditional rights and access were legally denied under forest and
40 wildlife conservation laws. The FRA recognises the right to individual land titles on land already
41 cleared as well as community forest rights such as collection of forest produce. Till November 2018, a
42 total of 64,328 community forest rights and a total of 17,040,343 individual land titles had been
43 approved and granted up to the end of 2017. Current concerns on policy and implementation gaps are
44 about strengths and pitfalls of decentralisation, identifying genuine right holders, verification of land
45 rights using technology and best practices, and curbing illegal claims (Sarap et al. 2013; Reddy et al.
46 2011; Aggarwal 2011; Ramnath 2008; Ministry of Environment and Forests and Ministry and Tribal
47 Affairs, Government of India 2010).

1 As per the FRA, the forest rights shall be conferred free of all encumbrances and procedural
2 requirements. Furthermore, without implementation of the provision of FRA on getting the informed
3 consent of local communities for both diversion of community forest land as well as for reforestation,
4 it poses legal and administrative hurdles in using existing forest land for implementation of India's
5 ambitious Green India Mission that aims to respond to climate change by a combination of adaptation
6 and mitigation measures in the forestry sector. It aims to increase forest/tree cover to the extent of 5
7 million hectares (Mha) and improve quality of forest/tree cover on another 5 Mha of forest/non-forest
8 lands and support forest based livelihoods of 3 million families and generate co-benefits through
9 ecosystem services (Government of India).

10 Thus, the community forest land recognised under FRA can be used for the purpose of Compensatory
11 Afforestation or restoration under REDD+ only with informed consent of the communities and a
12 decentralised mechanism for using CAMPA funds. India's forest and forest restoration can potentially
13 move away from a top-down carbon centric model with the effective participation of local
14 communities (Vijge and Gupta 2014; Murthy et al. 2018a).

15 India has also experimented with the world's first national inter-governmental ecological fiscal
16 transfer (EFT) from central to local and state government to reward them for retaining forest cover.
17 In 2014, India's 14th Finance Commission added forest cover to the formula that determines the
18 amount of tax revenue the central government distributes annually to each of India's 29 states. It is
19 estimated that in four years it would have distributed USD 6.9–12 billion per year to states in
20 proportion to their 2013 forest cover, amounting to around USD 174– 303 per hectare of forest per
21 year (Busch and Mukherjee 2017). State governments in India now have a sizeable fiscal incentive
22 based on extent of forest cover at the time of policy implementation contributing to the achievement
23 of India's climate mitigation and forest conservation goals. India's tax revenue distribution reform has
24 created the world's first EFTs for forest conservation, and a potential model for other countries.
25 However, it is to be noted that EFT is calculated based on a one-time estimate of forest cover prior to
26 policy implementation, hence does not incentivise ongoing protection and this is a policy gap. It's
27 still too early but its impact on trends in forest cover in the future and its ability to conserve forests
28 without other investments and policy instruments is promising but untested (Busch and Mukherjee
29 2017; Busch 2018).

30 In order to build on the new promising policy developments on forest rights and fiscal incentives for
31 forest conservation in India, incentivising ongoing protection, further investments in monitoring
32 (Busch 2018), decentralisation (Somanathan et al. 2009) and promotion of diverse non-agricultural
33 forest and range land based livelihoods (e.g., sustainable non-timber forest product extraction,
34 regulated pastures, carbon credits for forest regeneration on marginal agriculture land and ecotourism
35 revenues) as part of individual and community forest tenure and rights are ongoing concerns.
36 Decentralised sharing of CAMPA funds between government and local communities for forest
37 restoration as originally suggested and filling in implementation gaps could help reconcile climate
38 change mitigation through forest conservation, REDD+ and environmental justice (Vijge and Gupta
39 2014; Temper and Martinez-Alier 2013; Badola et al. 2013; Sun and Chaturvedi 2016; Murthy et al.
40 2018b; Chopra 2017; Ministry of Environment and Forests and Ministry and Tribal Affairs,
41 Government of India 2010).

7.4.7. Economic and financial instruments for adaptation, mitigation, and land

44 There is an urgent need to increase the volume of climate financing and bridge the gap between global
45 adaptation needs and available funds (*medium confidence*) (Valérie Masson-Delmotte et al. 2018;
46 Kissinger et al. 2019; Chambwera and Heal 2014), especially in relation to agriculture (FAO 2010).
47 The land sector offers the potential to balance the synergies between mitigation and adaptation

1 (Locatelli et al. 2016) (although context and unavailability of data sets makes cost comparisons
2 between mitigation and adaptation difficult (UNFCCC 2018b)). Estimates of adaptation costs range
3 from USD 140 to 300 billion by 2030, and between USD 280 and 500 billion by 2050; (UNEP 2016).
4 These figures vary according to methodologies and approaches (de Bruin et al. 2009; IPCC 2014
5 2014; Organization for Economic Cooperation and Development 2008; Nordhaus 1999; UNFCCC
6 2007; Plambeck et al. 1997).

7 **7.4.7.1. Financing mechanisms for land mitigation and adaptation**

8 A startling array of diverse and fragmented climate finance sources exist: more than 50 international
9 public funds, 60 carbon markets, 6000 private equity funds, 99 multilateral and bilateral climate funds
10 (Samuwai and Hills 2018). Most public finance for developing countries flows through bilateral and
11 multilateral institutions such as the World Bank, the International Monetary Fund, International
12 Finance Corporation, regional development banks, as well as specialised multilateral institutions such
13 as the Global Environmental Fund, and the EU Solidarity Fund. Some governments have established
14 state investment banks (SIBs) to close the financing gap, including the UK (Green Investment Bank),
15 Australia (Clean Energy Finance Corporation) and in Germany (Kreditanstalt für Wiederaufbau) the
16 Development Bank has been involved in supporting low-carbon finance (Geddes et al. 2018). The
17 Green Climate Fund (GCF) now offers additional finance, but is still a new institution with policy
18 gaps, a lengthy and cumbersome process related to approval (Brechtin and Espinoza 2017; Khan and
19 Roberts 2013; Mathy and Blanchard 2016), and challenges with adequate and sustained funding
20 (Schalatek and Nakhoda 2013). Private adaptation finance exists, but is difficult to define, track, and
21 coordinate (Nakhoda et al. 2016).

22 The amount of funding dedicated to agriculture, land degradation or desertification is very small
23 compared to total climate finance (FAO 2010). Funding for agriculture is accessed through the
24 smaller adaptation funds (rather than mitigation) (Lobell et al. 2013). Focusing on synergies, between
25 mitigation, adaptation, and increased productivity, such as through Climate Smart Agriculture
26 (CSA)(see 7.5.6), (Lipper et al. 2014b), may leverage greater financial resources (Suckall et al. 2015;
27 Locatelli et al. 2016). Payments for Ecosystem Services (see 7.4.6) are another emerging area to
28 encourage environmentally desirable practices, although they need to be carefully designed to be
29 effective (Engel and Muller 2016).

30 The UNCCD established the Land Degradation Neutrality Fund (LDN Fund) to mobilise finance and
31 scale up land restoration and sustainable business models on restored land to achieve the target of a
32 land degradation neutral world (SDG target 15.3) by 2030. The LDN Fund generates revenues from
33 sustainable use of natural resources, creating green job opportunities, sequestering CO₂, and
34 increasing food and water security (Cowie et al. 2018a; Akhtar-Schuster et al. 2017). The fund
35 leverages public money to raise private capital for sustainable land management and land restoration
36 projects (Quatrini and Crossman 2018; Stavi and Lal 2015). Many small-scale projects are
37 demonstrating that sustainable landscape management (see 7.6.3) is key to achieving LDN, and it is
38 also more financially viable in the long term than the unsustainable alternative (Tóth et al. 2018; Kust
39 et al. 2017).

40 **7.4.7.2. Instruments to manage the financial impacts of climate and land change** 41 **disruption**

42 Comprehensive risk management (see 7.4.3.1) designs a portfolio of instruments which are used
43 across a continuum of preemptive, planning and assessment, and contingency measures in order to
44 bolster resilience (Cummins and Weiss 2016) and address limitations of any one instrument
45 (Surminski 2016; Surminski et al. 2016; Linnerooth-bayer et al. 2019). Instruments designed and
46 applied in isolation have shown short-term rather than sustained intended impacts (Vincent et al.
47 2018). Risk assessments limited to events and impacts on particular asset classes or sectors can
48 misinform policy and drive misallocation of funding (Gallina et al. 2016; Jongman et al. 2014).

49 Comprehensive risk assessment combined with risk layering approaches that assign different
50 instruments to different magnitude and frequency of events, have better potential to provide stability

1 to societies facing disruption (Mechler et al. 2014; Surminski et al. 2016). Governments and citizens
2 define limits of what they consider acceptable risks, risks for which market or other solutions can be
3 developed and catastrophic risks that require additional public protection and intervention. Different
4 financial tools may be used for these different categories of risk or phases of the risk cycle
5 (preparedness, relief, recovery, reconstruction).

6 In order to protect lives and livelihoods early action is critical, including a coordinated plan for action
7 agreed in advance, a fast, evidence-based decision-making process, and contingency financing to
8 ensure that the plan can be implemented (Clarke and Dercon 2016a). Forecast-based finance
9 mechanisms incorporate these principles, using climate or other indicators to trigger funding and
10 action prior to a shock (Wilkinson 2018). Forecast-based mechanisms can be linked with social
11 protection systems by providing contingent scaled-up finance quickly to vulnerable populations
12 following disasters, enhancing scalability, timeliness, predictability and adequacy of social protection
13 benefits (Wilkinson 2018; Costella et al. 2017b; World Food Programme 2018).

14 Measures in advance of risks set aside resources before negative impacts related to adverse weather,
15 climatic stressors, and land changes occur. These tools are frequently applied in extreme event, rapid
16 onset contexts. These measures are the main instruments for reducing fatalities and limiting damage
17 from extreme climate and land change events (Surminski et al. 2016). Finance tools in advance of risk
18 include insurance (macro, meso, micro), green bonds, and forecast based finance (Hunzai et al. 2018).

19 There is *high confidence* that insurance approaches which are designed to effectively reduce and
20 communicate risks to the public and beneficiaries, designed to reduce risk and foster appropriate
21 adaptive responses, and provide value in risk transfer, improve economic stability and social
22 outcomes in both higher and lower income contexts (Kunreuther and Lyster 2016; Outreville
23 2011b)(Surminski et al. 2016; Kousky et al. 2018b), bolster food security, helping keep children in
24 school, and helping safeguard the ability of low income households to pay for essentials like
25 medicines (Shiferaw et al. 2014; Hallegatte et al. 2017).

26 Low income households show demand for affordable risk transfer tools, but demand is constrained by
27 liquidity, lack of assets, financial and insurance literacy, or proof of identity required by institutions in
28 the formal sector (Eling et al. 2014; Cole 2015; Cole et al. 2013; Ismail et al. 2017). Microinsurance
29 participation takes many forms including through mobile banking (Eastern Africa, Bangladesh),
30 linked with social protection or other social stabilisation programs (Ethiopia, Pakistan, India), through
31 flood or drought protection schemes (Indonesia, the Philippines, the Caribbean, and Latin America),
32 often in the form of weather index insurance. Insurance faces challenges around low public
33 awareness of how insurance works, risk, low capacity in financial systems to administer insurance,
34 data deficits, and market imperfections (Mechler et al. 2014; Feyen et al. 2011; Gallagher 2014;
35 Kleindorfer et al. 2012; Lazo et al.; Meyer and Priess 2014; Millo 2016).

36 Countries also request grant assistance, and contingency debt finance that includes dedicated funds,
37 set aside for unpredictable climate-related disasters, household savings, loans with “catastrophe risk
38 deferred drawdown option” (CATDDO) (which allows countries to divert loans from development
39 objectives such as health, education, and infrastructure to make immediate disbursement of funds in
40 the event of a disaster) (Kousky and Cooke 2012; Clarke and Dercon 2016b). Contingency finance is
41 suited to manage frequently occurring, low-impact events (Campillo et al. 2017; Mahul and
42 Ghesquiere 2010; Roberts 2017) and may be linked with social protection systems. These instruments
43 are limited by uncertainty surrounding the size of contingency fund reserves, given unpredictable
44 climate disasters (Roberts 2017) and lack of borrowing capacity of a country (such as small island
45 states) (Mahul and Ghesquiere 2010).

46 In part because of its link with debt burden, contingency, or post event finance can disrupt
47 development and is not suitable for higher consequence events and processes such as weather
48 extremes or structural changes associated with climate and land change. Post event finance of
49 negative impacts such as sea level rise, soil salinisation, depletion of groundwater, and widespread
50 land degradation is likely to become infeasible for multiple, high cost events and processes. There is
51 *high confidence* post-extreme event assistance may face more severe limitations given impacts of

1 climate change (Linnerooth-bayer et al. 2019; Surminski et al. 2016; Deryugina 2013; Dillon et al.
2 2014; Clarke 2016; Shreve and Kelman 2014; Von Peter et al. 2012).

3 In a catastrophe risk pool, multiple countries in a region pool risks in a diversified portfolio. Examples
4 include Africa Risk Capacity (ARC), the Caribbean Catastrophe Risk Insurance Facility (CCRIF), and
5 the Pacific Catastrophe Risk Assessment and Financing Initiative (PCRAFI) (Bresch et al. 2017;
6 Iyahen and Syroka 2018). ARC payouts have been used to assist over 2.1 million food insecure
7 people and provide over 900,000 cattle with subsidised feed in the affected countries (Iyahen and
8 Syroka 2018). ARC has also developed the Extreme Climate Facility, which is designed to
9 complement existing bilateral, multilateral and private sources of finance to enable proactive
10 adaptation (Vincent et al. 2018). It provides beneficiaries the opportunity to increase their benefit by
11 reducing exposure to risk through adaptation and risk reduction measures, thus side-stepping “moral
12 hazard” problems sometimes associated with traditional insurance.

13 Governments pay coupon interest when purchasing catastrophe (CAT) bonds from private or
14 corporate investors. In the case of the pre-defined catastrophe, the requirement to pay the coupon
15 interest or repay the principal may be deferred or forgiven (Nguyen and Lindenmeier 2014). CAT
16 bonds are typically short-term instruments (3–5 years) and the payout is triggered once a particular
17 threshold of disaster/damage is passed (Härdle and Cabrera 2010; Campillo et al. 2017; Estrin and
18 Tan 2016; Hermann, A., Koferl, P., Mairhofer 2016; Michel-Kerjan 2011; Roberts 2017). The
19 primary advantage of CAT bonds is their ability to quickly disburse money in the event of a
20 catastrophe (Estrin and Tan 2016). Green bonds, social impact bonds, and resilience bonds are other
21 instruments that can be used to fund land based interventions. However, there are significant barriers
22 for developing country governments to enter into the bond market: lack of familiarity with the
23 instruments; lack of capacity and resources to deal with complex legal arrangements; limited or non-
24 existent data and modelling of disaster exposure; and other political disincentives linked to insurance.
25 For these reasons the utility and application of bonds is currently largely limited to higher-income
26 developing countries (Campillo et al. 2017; Le Quesne 2017).

27 **7.4.7.3. Innovative financing approaches for transition to low carbon economies**

28 Traditional financing mechanisms have not been sufficient and thereby leave a gap in facilitating a
29 rapid transition to a low carbon economy or building resilience (Geddes et al. 2018). More recently
30 there have been developments in more innovative mechanisms including crowdfunding (Lam and
31 Law 2016), often supported by national governments (in the U.K. through regulatory and tax
32 support)(Owen et al. 2018). Crowdfunding has no financial intermediaries and thus low transaction
33 costs, and the projects have a greater degree of independence than bank or institution funding (Miller
34 et al. 2018). Other examples of innovative mechanisms are community shares for local projects, such
35 as renewable energy (Holstenkamp and Kahla 2016), or Corporate Power Purchase Agreements
36 (PPAs) used by companies such as Google and Apple to purchase renewable energy directly or
37 virtually from developers (Miller et al. 2018). Investing companies benefit from avoiding
38 unpredictable price fluctuations as well as increasing their environmental credentials. A second
39 example is auctioned price floors, or subsidies that offer a guaranteed price for future emission
40 reductions, currently being trialled in developing countries, by the World Bank Group, known as the
41 Pilot Auction Facility (PAF) (Bodnar et al. 2018). Price floors can maximise the climate impact per
42 public dollar while incentivising private investment in low-carbon technologies, and ideally would be
43 implemented in conjunction with complementary policies such as carbon pricing.

44 In order for climate finance to be as effective and efficient as possible, cooperation between private,
45 public and third sectors (e.g., NGOs, cooperatives, community groups) is more likely to create an
46 enabling environment for innovation (Owen et al. 2018). While innovative private sector approaches
47 are making significant progress, the existence of a stable policy environment that provides certainty
48 and incentives for long term private investment is critical.

49 **7.4.8. Enabling effective policy instruments – Policy Portfolio Coherence**

50 An enabling environment for policy effectiveness includes: 1) the development of comprehensive
51 policies, strategies and programs (section 7.4); 2) human and financial resources that ensure policies,

1 programs and legislation are translated into action; 3) decision making that draws on evidence
2 generated from functional information systems that make it possible to monitor trends; track and map
3 actions; and assess impact in a manner that is timely and comprehensive (see 7.5); 4) governance
4 coordination mechanisms and partnerships; and 5) a long term perspective in terms of response
5 options, monitoring, and maintenance (see 7.6) (FAO 2017a).

6 A comprehensive consideration of policy portfolios achieves sustainable land and climate
7 management (*medium confidence*) (Mobarak and Rosenzweig 2013; Stavropoulou et al. 2017)
8 (Jeffrey et al. 2017) (Howlett and Rayner 2013) (Aalto et al. 2017; Brander and Keith 2015; Williams
9 and Abatzoglou 2016) (Linnerooth-Bayer and Hochrainer-Stigler 2015) (FAO 2017b; Bierbaum and
10 Cowie 2018). Supporting the study of enabling environments, the study of policy mixes has emerged
11 in the last decade in regards to the mix or set of instruments that interact together and are aimed at
12 achieving policy objectives in a dynamic setting (Reichardt et al. 2015). The study of policy mixes
13 includes studying the ultimate objectives of a policy mix (such as biodiversity (Ring and Schröter-
14 Schlaack 2011)), the interaction of policy instruments within the mix (including climate change
15 mitigation and energy (del Río and Cerdá 2017)) (see Trade-offs and Synergies, 7.5.6), and the
16 dynamic nature of the policy mix (Kern and Howlett 2009)).

17 Studying policy mixes allows for a consideration of policy coherence which is broader than the study
18 of discrete policy instruments in rigidly defined sectors, but entails studying policy in relation to the
19 links and dependencies among problems and issues (FAO 2017b). Consideration of policy coherence
20 is a new approach rejecting simplistic solutions, but acknowledging inherently complex processes
21 involving collective consideration of public and private actors in relation to policy analysis (FAO
22 2017b). A coherent, consistent mix of policy instruments can solve complex policy problems
23 (Howlett and Rayner 2013) as it involves lateral, integrative, and holistic thinking in defining and
24 solving problems (FAO 2017b). Such a consideration of policy coherence is required to achieve
25 sustainable development (FAO 2017b; Bierbaum and Cowie 2018). Considerations of policy
26 coherence potentially addresses three sets of challenges: challenges that exist with assessing multiple
27 hazards and sectors (Aalto et al. 2017; Brander and Keith 2015; Williams and Abatzoglou 2016);
28 challenges in mainstreaming adaptation and risk management into on-going development planning
29 and decision making (Linnerooth-Bayer and Hochrainer-Stigler 2015); challenges in scaling up
30 community and ecosystem based initiatives in countries overly focused on sectors, instead of
31 sustainable use of biodiversity and ecosystem services (Reid 2016). There is a gap in integrated
32 consideration of adaptation, mitigation, climate change policy and development. A study in Indonesia
33 found while internal policy coherence between mitigation and adaptation is increasing, external policy
34 coherence between climate change policy and development objectives is still required (Di Gregorio et
35 al. 2017).

36 There is *medium evidence and high agreement* that a suite of agricultural business risk programs
37 (which would include crop insurance and income stability programs) increase farm financial
38 performance, reduce risk, and also reinforce incentives to adopt stewardship practices (beneficial
39 management practices) improving the environment (Jeffrey et al. 2017). Consideration of the portfolio
40 of instruments responding to climate change and its associated risks, and the interaction of policy
41 instruments, improve agricultural producer livelihoods (Hurlbert 2018b). In relation to hazards, or
42 climate related extremes (7.4.3), the policy mix has been found to be a key determinant of the
43 adaptive capacity of agricultural producers. In relation to drought, the mix of policy instruments
44 including crop insurance, sustainable land management practices, bankruptcy and insolvency, co-
45 management of community in water and disaster planning, and water infrastructure programmes are
46 effective at responding to drought (Hurlbert 2018b; Hurlbert and Mussetta 2016; Hurlbert and Pittman
47 2014; Hurlbert and Montana 2015; Hurlbert 2015a) (Hurlbert and Gupta 2018). Similarly in relation
48 to flood, the mix of policy instruments including flood zone mapping, land use planning, flood zone

1 building restrictions, business and crop insurance, disaster assistance payments, preventative
2 instruments including environmental farm planning (including soil and water management (see
3 Chapter 6)) and farm infrastructure projects, and recovery from debilitating flood losses ultimately
4 through bankruptcy are effective at responding to flood (Hurlbert 2018a)(see 7.6.3 Case Study Flood
5 and Flood Security).

6 In respect of land conservation and management goals, consideration of differing strengths and
7 weakness of instruments is necessary. While direct regulation may secure effective minimum
8 standards of biodiversity conservation and critical ecosystem service provision, economic instruments
9 may achieve reduced compliance costs as costs are borne by policy addressees (Rogge and Reichardt
10 2016). In relation to GHG emissions and climate mitigation a comprehensive mix of instruments
11 targeted at emissions reductions, learning, and research and development is effective (*high*
12 *confidence*) (Fischer and Newell 2008). The policy coherence between climate policy and public
13 finance is critical in ensuring the efficiency, effectiveness and equity of mitigation policy, and
14 ultimately to make stringent mitigation policy more feasible (Siegmeier et al. 2018). Recycling
15 carbon tax revenue to support clean energy technologies can decrease losses from unilateral carbon
16 mitigation targets with complementary technology policies (Corradini et al. 2018).

17 When evaluating a new policy instrument, its design in relation to achieving an environmental goal or
18 solving a land and climate change issue, includes consideration of how the new instrument will
19 interact with existing instruments operating at multiple levels (international, regional, national, sub-
20 national, and local) (Ring and Schröter-Schlaack 2011)(see 7.4.1).

21

22 **7.4.9. Barriers to Implementing Policy Responses**

23 There are barriers to implementing the policy instruments that arise in response to the risks from
24 climate-land interactions. Such barriers to climate action help determine the degree to which society
25 can achieve its sustainable development objectives (Dow et al. 2013; Langholtz et al. 2014; Klein et
26 al. 2015). However, some policies can also be seen as being designed specifically to overcome
27 barriers, while in some cases policies may actually create or strengthen barriers to climate action
28 (Foudi and Erdlenbruch 2012; Linnerooth-Bayer and Hochrainer-Stigler 2015). The concept of
29 barriers to climate action is used here in a sense close to that of “soft limits” to adaptation (Klein, et
30 al. 2014). “Hard limits” by contrast are seen as primarily biophysical. Predicted changes in the key
31 factors of crop growth and productivity—temperature, water, and soil quality— are expected to pose
32 limits to adaptation in ways that affect the world’s population to get enough food in the future (Altieri
33 et al. 2015; Altieri and Nicholls 2017).

34 This section assesses research on barriers specific to policy implementation in adaptation and
35 mitigation respectively, then addresses the cross-cutting issue of inequality as a barrier to climate
36 action, including the particular cases of elite capture and corruption, before assessing how policies on
37 climate and land can be used to overcome barriers.

38 **7.4.9.1. Barriers to Adaptation**

39 There are human, social, economic, and institutional barriers to adaptation to land-climate challenges
40 as described in Tabel 7.4 (*medium evidence, high agreement*). Considerable literature exists around
41 changing behaviours through response options targeting social and cultural barriers (Rosin 2013;
42 Eakin; Marshall et al. 2012) (See Chapter 6 Value chain interventions).

1

Table 7.4 Soft Barriers and Limits to Adaptation

| Category | Description | References |
|---------------|---|---|
| Human | Cognitive and behavioural obstacles. Lack of knowledge and information. | (Hornsey et al. 2016; Prokopy et al. 2015) (Wreford et al. 2017) |
| Social | Undermined participation in decision making and social equity | (Burton et al. 2008) (Laube et al. 2012) |
| Economic | Market failures and missing markets, transaction costs and political economy, ethical and distributional issues. Perverse incentives. Lack of domestic funds, inability to access international funds | (Chambwera et al. 2014b) (Wreford et al. 2017) (RocheCouste et al. 2015; Baumgart-Getz et al. 2012) |
| Institutional | Mal-coordination of policies and response options, unclear responsibility of actors and leadership, misuse of power, all reducing social learning. Government failures. Path dependent institutions. | (Oberlack 2017) (Sánchez et al. 2016; Greiner and Gregg 2011) |
| Technological | Systems of mixed crop and livestock. Polycultures. | (Nalau and Handmer 2015) |

2

3 Since AR5 research examining the role of governance, institutions and in particular policy
4 instruments, in creating or overcoming barriers to adaptation to land and climate change in the land
5 use sector is emerging (Foudi and Erdlenbruch 2012; Linnerooth-Bayer and Hochrainer-Stigler 2015).
6 Evidence shows that understanding the local context and targeted approaches are generally most
7 successful (Rauken et al. 2014). Understanding the nature of constraints to adaptation is critical in
8 determining how barriers may be overcome. Formal institutions (rules, laws, policies) and informal
9 institutions (social and cultural norms and shared understandings) can be barriers and enablers of
10 climate adaptation (Jantarasami et al. 2010). Governments play a key role in intervening and
11 confronting existing barriers by changing legislation, adopting policy instruments, providing
12 additional resources, and building institutions and knowledge exchange (Ford and Pearce 2010;
13 Measham et al. 2011; Mozumder et al. 2011; Storbjörk 2010). Understanding institutional barriers is
14 important in addressing barriers (*high confidence*). Institutional barriers may exist due to the path-
15 dependent nature of institutions governing natural resources and public good, bureaucratic structures
16 that undermine horizontal and vertical integration (see 7.6.2), and lack of policy coherence (see 7.4.8).

17 Governments play a key role in intervening and confronting existing barriers by changing legislation,
18 adopting policy instruments, providing additional resources, and building institutions and knowledge
19 exchange (Ford and Pearce 2010; Measham et al. 2011; Mozumder et al. 2011; Storbjörk 2010).
20 Understanding institutional barriers is important in addressing barriers (*high confidence, robust
21 evidence*). Institutional barriers may exist due to the path-dependent nature of institutions governing
22 natural resources and public good, bureaucratic structures that undermine horizontal and vertical
23 integration (see 7.6.2), and lack of policy coherence (see 7.4.8). Governments play a key role in
24 intervening and confronting existing barriers by changing legislation, adopting policy instruments,
25 providing additional resources, and building institutions and knowledge exchange (Ford and Pearce
26 2010; Measham et al. 2011; Mozumder et al. 2011; Storbjörk 2010). Understanding institutional
27 barriers is important in addressing barriers (*high confidence, robust evidence*). Institutional barriers
28 may exist due to the path-dependent nature of institutions governing natural resources and public
29 good, bureaucratic structures that undermine horizontal and vertical integration (see 7.6.2), and lack
30 of policy coherence (see 7.4.8).

7.4.9.2. Barriers to land based climate mitigation

Barriers to land based mitigation relate to full understanding of the permanence of carbon sequestration in soils or terrestrial biomass, the additionality of this storage, its impact on production and production shifts to other regions, measurement and monitoring systems and costs (Smith et al. 2007). Agricultural producers are more willing to expand mitigation measures already employed (including efficient and effective management of fertiliser including manure and slurry) and less favourable to those not employed such as using dietary additives, adopting genetically improved animals, or covering slurry tanks and lagoons (Feliciano et al. 2014). Barriers identified in land based mitigation include physical environmental constraints including lack of information, education, and suitability for size and location of farm. For instance precision agriculture is not viewed as efficient in small scale farming (Feliciano et al. 2014).

Property rights may be a barrier when there is no clear single party land ownership to implement and manage changes (Smith et al. 2007). In forestry, tenure arrangements may not distribute obligations and incentives for carbon sequestration effectively between public management agencies and private agents with forest licenses. Including carbon in tenure and expanding the duration of tenure may provide stronger incentive for tenure holders to manage carbon as well as timber values (Williamson and Nelson 2017). Effective policy will require answers as to the current status of agriculture in regard to GHG emissions, the degree that emissions are to change, the best pathway to achieve the change, and an ability to know when the target level of change is achieved (Smith et al. 2007). Forest governance may not have the structure to advance mitigation and adaptation. Currently top down traditional modes do not have the flexibility or responsiveness to deal with the complex, dynamic, spatially diverse, and uncertain features of climate change (Timberlake and Schultz 2017; Williamson and Nelson 2017).

In respect of forest mitigation, two main institutional barriers have been found to predominate. First forest management institutions do not consider climate change to the degree necessary for enabling effective climate response and do not link adaptation and mitigation; Second, institutional barriers exist if institutions are not forward looking, do not enable collaborative adaptive management, promote flexible approaches that are reversible as new information becomes available, promote learning and allow for diversity of approaches that can be tailored to different local circumstances (Williamson and Nelson 2017).

Land-based climate mitigation through expansions and enhancements in agriculture, forestry and bioenergy has great potential but also poses great risks and its success will therefore require improved land use planning, strong governance frameworks and coherent and consistent policies. “Progressive developments in governance of land and modernisation of agriculture and livestock and effective sustainability frameworks can help realise large parts of the technical bioenergy potential with low associated GHG emissions”(Smith et al. 2014b, p. 97).

7.4.9.3. Inequality

There is *medium evidence and high agreement* that one of the greatest challenges for land based adaptation and sustainable land management is posed by inequalities that influence vulnerability and coping and adaptive capacity - including age, gender, wealth, knowledge, access to resources and power (Kunreuther et al. 2014; IPCC 2012; Olsson et al. 2014). Gender is the dimension of inequality that has been the focus of most research while research demonstrating differential impacts, vulnerability and adaptive capacity based on age, ethnicity and indigeneity is less well developed (Olsson et al. 2015a). Cross-Chapter Box 11 sets out both the contribution of gender relations to differential vulnerability and available policy instruments for greater gender inclusivity.

One response to the vulnerability of poor people and other categories differentially affected is effective and reliable social safety nets (Jones and Hiller 2017). Social protection coverage is low across the world and informal support systems continue to be the key means of protection for a

1 majority of the rural poor and vulnerable (Stavropoulou et al. 2017)(See 7.4.2). However, there is a
2 gap in knowledge in understanding both positive and negative synergies between formal and informal
3 systems of social protection and how local support institutions might be used to implement more
4 formal forms of social protection (Stavropoulou et al. 2017).

5 **7.4.9.4. Corruption and elite capture**

6 Inequalities of wealth and power can allow processes of corruption and elite capture which can affect
7 both adaptation and mitigation actions, at levels from the local to the global, that in turn risk creating
8 inequitable or unjust outcomes (Sovacool 2018) (*limited evidence, medium agreement*). This includes
9 risks of corruption in REDD+ processes (Sheng et al. 2016; Williams and Dupuy 2018) and of
10 corruption or elite capture in broader forest governance (Sundström 2016; Persha and Andersson
11 2014), as well as elite capture of benefits from planned adaptation at a local level (Sovacool 2018).

12 Peer-reviewed empirical studies that focus on corruption in climate finance and climate interventions,
13 particularly at a local level, are rare, due in part to the obvious difficulties of researching illegal and
14 clandestine activity (Fadairo et al. 2017). At the country level, historical levels of corruption are
15 shown to affect current climate policies and global cooperation (Fredriksson and Neumayer 2016).
16 Brown (2010) sees three likely inlets of corruption into REDD: in the setting of forest baselines, the
17 reconciliation of project and natural credits, and the implementation of control of illegal logging. The
18 transnational and north-south dimensions of corruption are highlighted by debates on which US
19 legislative instruments (e.g., the Lacey Act, the Foreign Corrupt Practices Act) could be used to
20 prosecute the northern corporations that are involved in illegal logging (Gordon 2016; Waite 2011).

21 Fadairo et al. (2017) carried out a structured survey of perceptions of households in forest-edge
22 communities served by REDD+, as well as those of local officials, in south eastern Nigeria. They
23 report high rates of agreement that allocation of carbon rights is opaque and uncertain, distribution of
24 benefits is untimely, uncertain and unpredictable, and the REDD+ decision-making process is
25 vulnerable to political interference that benefits powerful individuals. Only 35% of respondents had
26 an overall perception of transparency in REDD+ process as “good”. Of eight institutional processes or
27 facilities previously identified by the Government of Nigeria and international agencies as indicators
28 of commitment to transparent and equitable governance, only three were evident in the local REDD+
29 office as “very functional” or “fairly functional”.

30 At the local level, the risks of corruption and elite capture of the benefits of climate action are high in
31 decentralised regimes (Persha and Andersson 2014). (Rahman 2018) discusses elicitation of bribes
32 (by local-level government staff) and extortion (by criminals) to allow poor rural people to gather
33 forest products. The results are a general undermining of households’ adaptive capacity and perverse
34 incentives to over-exploit forests once bribes have been paid, leading to over-extraction and
35 biodiversity loss. Where there are pre-existing inequalities and conflict, participation processes need
36 careful management and firm external agency to achieve genuine transformation and avoid elite
37 capture (Rigon 2014). An illustration of the range of types of elite capture is given by Sovacool
38 (2018) for adaptation initiatives including coastal afforestation, combining document review and key
39 informant interviews in Bangladesh, with an analytical approach from political ecology. Four
40 processes are discussed: enclosure, including land grabbing and preventing the poor establishing new
41 land rights; exclusion of the poor from decision-making over adaptation; encroachment on the
42 resources of the poor by new adaptation infrastructure; and entrenchment of community
43 disempowerment through patronage. The article notes that observing these processes does not imply
44 they are always present, nor that adaptation efforts should be abandoned.

45 **7.4.9.5. Overcoming Barriers**

46 Policy instruments that strengthen agricultural producer assets or capitals reduce vulnerability and
47 overcome barriers to adaptation (Hurlbert 2018b, 2015b). Additional factors like formal education
48 and knowledge of traditional farming systems, secure tenure rights, access to electricity and social

1 institutions in rice-farming areas of Bangladesh have played a positive role in reducing adaptation
2 barriers (Alam 2015). A review of over 168 publications over 15 years about adaptation of water
3 resources for irrigation in Europe found the highest potential for action is in improving adaptive
4 capacity and responding to changes in water demands, in conjunction with alterations in current water
5 policy, farm extension training, and viable financial instruments (Iglesias and Garrote 2015). Research
6 on the Great Barrier Reef, the Olifants River in Southern Africa, and fisheries in Europe, North
7 America, and the Antarctic Ocean, suggests the leading factors in harnessing the adaptive capacity of
8 ecosystems is to reduce human stressors by enabling actors to collaborate across diverse interests,
9 institutional settings, and sectors (Biggs et al. 2017; Schultz et al. 2015; Johnson and Becker 2015).
10 Fostering equity and participation are correlated with the efficacy of local adaptation to secure food
11 and livelihood security (Laube et al. 2012). In this chapter, the literature surrounding appropriate
12 policy instruments, decision making, and governance practices to overcome limits and barriers to
13 adaptation is proposed.

14 Incremental adaptation consists of actions where the central aim is to maintain the essence and
15 integrity of a system or process at a given site whereas transformational adaptation is adaptation that
16 changes the fundamental attributes of a system in response to climate and its effects; the former is
17 characterised as doing different things and the latter, doing things differently (Noble et al. 2014).
18 Transformational adaptation is necessary in situations where there are hard limits to adaptation or it is
19 desirable to address deficiencies in sustainability, adaptation, inclusive development and social equity
20 (Kates et al. 2012; Mapfumo et al. 2016). In other situations, incremental changes may be sufficient
21 (Hadarits et al. 2017).

22

23 **Cross-chapter Box 11: Gender in inclusive approaches to climate** 24 **change, land, and sustainable development**

25

26 Margot Hurlbert (Canada), Brigitte Baptiste (Colombia), Amber Fletcher (Canada), Marta Guadalupe
27 Rivera Ferre (Spain), Darshini Mahadevia (India), Katharine Vincent (United Kingdom)

28

29 Gender is a key axis of social inequality that intersects with other systems of power and
30 marginalisation—including “race”, culture, class/socioeconomic status, location, sexuality, and age—
31 to cause unequal experiences of climate change vulnerability and adaptive capacity. However, “policy
32 frameworks and strong institutions that align development, equity objectives, and climate have the
33 potential to deliver ‘triple-wins’” (Roy et al. 2018), including enhanced gender equality. Gender in
34 relation to this report is introduced in Chapter 1, referred to as a leverage point in women’s
35 participation in decisions relating to land desertification (3.6.3), land degradation (4.1.6), food
36 security (5.2.5.1), and enabling land and climate response options (6.1.2.2).

37

38 Focusing on ‘gender’ as a relational and contextual construct can help avoid homogenising “women”
39 as a uniformly and consistently vulnerable category (Arora-Jonsson 2011; Mersha and Van Laerhoven
40 2016; Ravera et al. 2016). There is *high agreement* that using a framework of intersectionality to
41 integrate gender into climate change research helps to recognise overlapping and interconnected
42 systems of power (Djoudi et al. 2016; Fletcher 2018; Kaijser and Kronsell 2014; Moosa and Tuana
43 2014; Thompson-Hall et al. 2016), which create particular inequitable experiences of climate change
44 vulnerability and adaptation. Through this framework, both commonalities and differences may be
45 found between the experiences of rural and urban women, or between women in high-income and
46 low-income countries, for example.

47

48 In rural areas, women generally experience greater vulnerability than men, albeit through different
49 pathways (Djoudi et al., 2016; Goh, 2012; Jost et al., 2016; Kakota, Nyariki, Mkwambisi, & Kogi-

1 Makau, 2011). In masculinised agricultural settings of Australia and Canada, for example, climate
2 adaptation can increase women's work on- and off-farm, but without increasing recognition for
3 women's undervalued contributions (Alston et al. 2018a; Fletcher and Knuttila 2016). A study in
4 rural Ethiopia found that male-headed households had access to a wider set of adaptation measures
5 than female-headed households (Mersha and Van Laerhoven 2016).

6
7 Due to engrained patriarchal social structures and gendered ideologies, women may face multiple
8 barriers to participation and decision-making in land-based adaptation and mitigation actions in
9 response to climate change (*high confidence*) (Alkire et al. 2013a; Quisumbing et al. 2014). These
10 barriers include: (i) disproportionate responsibility for unpaid domestic work, including care-giving
11 activities (Beuchelt and Badstue 2013) and provision of water and firewood (UNEP, 2016); (ii) risk
12 of violence in both public and private spheres, which restricts women's mobility for capacity-building
13 activities and productive work outside the home (Day et al., 2005; Jost et al., 2016; UNEP, 2016); (iii)
14 less access to credit and financing (Jost et al. 2016); (iv) lack of organisational social capital, which
15 may help in accessing credit (Carroll et al. 2012); (v) lack of ownership of productive assets and
16 resources (Kristjanson et al., 2014; Meinzen-Dick et al., 2010), including land. Constraints to land
17 access include not only state policies, but also customary laws (Bayisenge 2018) based on customary
18 norms and religion that determine women's rights (Namubiru-Mwaura 2014a).

19
20 Differential vulnerability to climate change is related to inequality in rights-based resource access,
21 established through formal and informal tenure systems. In only 37% of 161 developing and
22 developed countries do men and women have equal rights to use and control land, and in 59%
23 customary, traditional, and religious practices discriminate against women (OECD 2014), even if the
24 law formally grants equal rights. Women play a significant role in agriculture, food security and rural
25 economies globally, forming 43% of the agricultural labour force in developing countries (FAO,
26 IFAD, UNICEF, & WHO, 2018, p. 102), ranging from 25 % in Latin America (FAO, 2017, pp. 89) to
27 nearly 50% in Eastern Asia and Central and South Europe (FAO, 2017, p. 88) and 47% in sub-
28 Saharan Africa (FAO, 2017, pp. 88). Further, the share of women in agricultural employment has
29 been growing in all developing regions except East Asia and Southeast Asia (FAO, 2017, p. 88). At
30 the same time, women constitute less than 5% of landholders (with legal rights and/or use-rights
31 (Doss et al. 2018a) in North Africa and West Asia, about 15% in sub-Saharan Africa, 12% in
32 Southern and Southeastern Asia, 18% in Latin America and Caribbean (FAO 2011b, p. 25), 10% in
33 Bangladesh, 4% in Nigeria (FAO 2015c). Patriarchal structures and gender roles can also affect
34 women's control over land in developed countries (Carter 2017; Alston et al. 2018b). Thus,
35 longstanding gender inequality in land rights, security of tenure, and decision-making may constrict
36 women's adaptation options (Smucker and Wangui 2016).

37
38 Adaptation options related to land and climate (see Chapter 6) may produce environment and
39 development trade-offs as well as social conflicts (Hunsberger et al. 2017) and changes with gendered
40 implications. Women's strong presence in agriculture provides opportunity to bring gender
41 dimensions into climate change adaptation, particularly regarding food security (Glemarec 2017; Jost
42 et al. 2016; Doss et al. 2018b). Some studies point to a potentially emancipatory role played by
43 adaptation interventions and strategies, albeit with some limitations depending on context. For
44 example, in developing contexts, male out-migration may cause women in socially disadvantaged
45 groups to engage in new livelihood activities, thus challenging gendered roles (Djoudi and Brockhaus
46 2011; Alston 2006). Collective action and agency of women in farming households, including
47 widows, have led to prevention of crop failure, reduced workload, increased nutritional intake,
48 increased sustainable water management, diversified and increased income and improved strategic
49 planning (Andersson and Gabrielsson 2012). Women's waged labour can help stabilise income from
50 more land- and climate-dependent activities such as agriculture, hunting, or fishing (Alston et al.,
51 2018; Ford & Goldhar, 2012). However, in developed contexts like Australia, women's participation
52 in off-farm employment may exacerbate existing masculinisation of agriculture (Clarke and Alston
53 2017).

54
55 Literature suggests that land-based mitigation measures may lead to land alienation either through

1 market or appropriation (acquisition) by the government, interfere with traditional livelihoods in rural
2 areas, and lead to decline in women's livelihoods (Hunsberger et al. 2017). If land alienation is not
3 prevented, existing inequities and social exclusions may be reinforced (*medium agreement*)
4 (Mustalahti and Rakotonarivo 2014; Chomba et al. 2016; Poudyal et al. 2016). These activities also
5 can lead to land grabs, which remain a focal point for research and local activism (Borras Jr. et al.
6 2011; White et al. 2012; Lahiff 2015). Cumulative effects of land-based mitigation measures may put
7 families at risk of poverty. In certain contexts, they lead to increased conflicts. In conflict situations,
8 women are at risk of personal violence, including sexual violence (UNEP, 2016).

9 10 **Policy instruments for gender inclusive approaches to climate change, land, and sustainable** 11 **development**

12
13 Integrating, or mainstreaming, gender into land and climate change policy requires assessments of
14 gender-differentiated needs and priorities, selection of appropriate policy instruments to address
15 barriers to women's sustainable land management, and selection of gender indicators for monitoring
16 and assessment of policy (*medium confidence*) (Huyer et al. 2015a; Alston 2014). Important sex-
17 disaggregated data can be obtained at multiple levels, including the intra-household level (Seager
18 2014; Doss et al. 2018b), village- and plot-level information (Theriault et al. 2017a), and through
19 national surveys (Agarwal 2018a; Doss et al. 2015a). Gender-disaggregated data provides a basis for
20 selecting, monitoring and reassessing policy instruments that account for gender differentiated land
21 and climate change needs (*medium confidence*) (Rao 2017a; Arora-Jonsson 2014; Theriault et al.
22 2017b) (Doss et al. 2018b). While macro-level data can reveal ongoing gender trends in SLM,
23 contextual data are important for revealing intersectional aspects, such as the difference made by
24 family relations, socioeconomic status, or cultural practices about land use and control (Rao 2017a;
25 Arora-Jonsson 2014; Theriault et al. 2017b), as well as on security of land holding (Doss et al.
26 2018b). Indices such as the Women's Empowerment in Agriculture Index (Alkire et al. 2013b) may
27 provide useful guidelines for quantitative data collection on gender and SLM, while qualitative
28 studies can reveal the nature of agency and whether policies are likely to be accepted, or not, in the
29 context of local structures, meanings, and social relations (Rao 2017b).

30
31 Women's economic empowerment, decision-making power and voice is a necessity in SLM decisions
32 (Mello and Schmink 2017a; Theriault et al. 2017b). Policies that address barriers include: gender
33 considerations as qualifying criteria for funding programs or access to financing for initiatives;
34 government transfers to women under the auspices of anti-poverty programs; spending on health and
35 education; and subsidised credit for women (*medium confidence*) (Jagger and Pender 2006; Van
36 Koppen et al. 2013a; Theriault et al. 2017b; Agarwal 2018b). Training and extension for women to
37 facilitate sustainable practices is also important (Mello and Schmink 2017b; Theriault et al. 2017b).
38 Such training could be built into existing programs or structures, such as collective microenterprise
39 (Mello and Schmink 2017b). Huyer et al. (2015) suggest that information provision (e.g., information
40 about SLM) could be effectively dispersed through women's community-based organisations,
41 although not in such a way that it overwhelms these organisations or supersedes their existing
42 missions. SLM programs could also benefit from intentionally engaging men in gender-equality
43 training and efforts (Fletcher 2017), thus recognising the relationality of gender. Recognition of the
44 household level, including men's roles and power relations, can help avoid the de-contextualised and
45 individualistic portrayal of women as purely instrumental actors (Rao 2017b).

46
47 Technology, policy, and programs that exacerbate women's workloads or reinforce gender stereotypes
48 (MacGregor 2010; Huyer et al. 2015b), or which fail to recognise and value the contributions women
49 already make (Doss et al. 2018b), may further marginalise women. Accordingly, some studies have
50 described technological and labour interventions that can enhance sustainability while also decreasing
51 women's workloads; for example, Vent et al. (2017) described the system of rice intensification as
52 one such intervention. REDD+ initiatives need to be aligned with the SDGs to achieve
53 complementary synergies with gender dimensions.

54
55 Secure land title and/or land access/control for women increases sustainable land management by

1 increasing women's conservation efforts, increasing their productive and environmentally-beneficial
2 agricultural investments, such as willingness to engage in tree planting and sustainable soil
3 management (*high confidence*) as well as improving cash incomes (Higgins et al. 2018; Agarwal
4 2010; Namubiru-Mwaura 2014b; Doss et al. 2015b; Van Koppen et al. 2013b; Theriault et al. 2017b;
5 Jagger and Pender 2006). According FAO (2011b, p. 5), if women had the same access to productive
6 resources as men, the number of hungry people in the world could be reduced by 12-17%. Policies
7 promoting secure land title include legal reforms at multiple levels, including national laws on land
8 ownership, legal education, and legal aid for women on land ownership and access (Argawal 2018).
9 Policies to increase women's access to land could occur through three main avenues of land
10 acquisition: inheritance/family (Theriault et al. 2017b), state policy, and the market (Agarwal 2018).
11 Rao (2017) recommends framing land rights as entitlements rather than as instrumental means to
12 sustainability. This reframing may address persistent, pervasive gender inequalities (FAO 2015d).
13
14
15

16 **7.5. Decision-making for Climate Change and Land**

17 The risks posed by climate change generate considerable uncertainty and complexity for decision-
18 makers responsible for land use decisions (*robust evidence, high agreement*). Decision-makers
19 balance climate ambitions, encapsulated in the NDCs, with other SDGs, which will differ
20 considerably across different regions, sociocultural conditions and economic levels (Griggs et al.
21 2014). The interactions across SDGs also factor into decision-making processes (Nilsson et al.
22 2016b). The challenge is particularly acute in Least Developed Countries where a large share of the
23 population is vulnerable to climate change. Matching the structure of decision-making processes to
24 local needs while connecting to national strategies and international regimes is challenging (Nilsson
25 and Persson 2012). This section explores methods of decision-making to address the risks and inter-
26 linkages outlined in previous sections. As a result, this section outlines policy inter-linkages with
27 SDGs and NDCs, trade-offs and synergies in specific measures, possible challenges as well as
28 opportunities going forward.

29 Even in cases where uncertainty exists, there is *medium evidence and high agreement* in the literature
30 that it need not present a barrier to taking action, and there are growing methodological developments
31 and empirical applications to support decision-making. Progress has been made in identifying key
32 source of uncertainty and addressing them (Farber 2015; Lawrence et al. 2018; Bloemen et al. 2018).
33 Many of these approaches involve principles of robustness, diversity, flexibility, learning, or choice
34 editing (see 7.5.2).

35 Since the Fifth Assessment Report Chapter on Decision-making (Jones et al. 2014) considerable
36 advances have been made in decision making under uncertainty, both conceptually and in economics
37 (see 7.5.2), and in the social/qualitative research areas (see 7.5.3 and 7.5.4). In the land sector, the
38 degree of uncertainty varies and is particularly challenging for climate change adaptation decisions
39 (Hallegatte 2009; Wilby and Dessai 2010). Some types of agricultural production decisions can be
40 made in short time-frames as changes are observed, and will provide benefits in the current time
41 period (Dittrich et al. 2017).

42 **7.5.1. Formal and Informal decision-making**

43 Informal decision making facilitated by open platforms can solve problems in land and resource
44 management by allowing evolution and adaptation, and incorporation of local knowledge (*medium*
45 *confidence*) (Malogdos and Yujuico 2015a; Vandersypen et al. 2007). Formal centers of decision
46 making are those that follow fixed procedures (written down in statutes or moulded in an organisation

1 backed by the legal system) and structures (Onibon et al. 1999). Informal centers of decision making
2 are those following customary norms and habits based on conventions (Onibon et al. 1999) where
3 problems are ill-structured, complex problems (Waddock 2013).

4 **7.5.1.1. Formal Decision Making**

5 Formal decision making processes can occur at all levels including the global, regional, national and
6 sub-national levels (see 7.4.1). Formal decision support tools can be used, for example, by farmers, to
7 answer “what-if” questions as to how to respond to the effects of changing climate on soils, rainfall
8 and other conditions (Wenkel et al. 2013).

9 Optimal formal decision-making is based on realistic behaviour of actors, important in land-climate
10 systems, assessed through participatory approaches, stakeholder consultations and by incorporating
11 results from empirical analyses. Mathematical simulations and games (Lamarque et al. 2013),
12 behavioural models in land-based sectors (Brown et al. 2017), agent-based models (ABMs) and
13 micro-simulations are examples useful to decision-makers (Bishop et al. 2013). These decision
14 making tools are expanded on in 7.5.2.

15 There are different ways to incorporate local knowledge, informal institutions and other contextual
16 characteristics that capture non-deterministic elements, as well as social and cultural beliefs and
17 systems more generally, into formal decision making (see 7.6.4) (*medium evidence, medium*
18 *agreement*). Classic scientific methodologies now include participatory and interdisciplinary methods
19 and approaches (Jones et al. 2014). Consequently, this broader range of approaches may very well
20 capture informal and indigenous knowledge improving the participation of indigenous peoples in
21 decision-making processes and thereby promote their rights to self-determination (Malogdos and
22 Yujuico 2015b) (see Cross-Chapter Box 13: Indigenous and Local Knowledge in this chapter).

23 **7.5.1.2. Informal Decision Making**

24 Informal institutions have contributed to sustainable resources management (common pool resources)
25 through creating a suitable environment for decision-making. The role of informal institutions and
26 decision making can be particularly relevant for land use decisions and practices in rural areas in the
27 global south and north (Huisheng 2015). Understanding informal institutions is crucial for adapting
28 to climate change, advancing technological adaptation measures achieving comprehensive disaster
29 management and advancing collective decision making (Karim and Thiel 2017). Informal institutions
30 have been found to be a crucial entry point in dealing with vulnerability of communities and
31 exclusionary tendencies impacting marginalised and vulnerable people (Mubaya and Mafongoya
32 2017).

33 Many studies underline the role of local/informal traditional institutions in the management of natural
34 resources in different parts of the world (Yami et al. 2009; Zoogah et al. 2015; Bratton 2007; Mowo et
35 al. 2013; Grzymala-Busse 2010). Traditional systems include: traditional silvo-pastoral management
36 (Iran), management of rangeland resources (South Africa), natural resource management (Ethiopia,
37 Tanzania, Bangladesh) communal grazing land management (Ethiopia) and management of conflict
38 over natural resources (Siddig et al. 2007; Yami et al. 2011; Valipour et al. 2014; Bennett 2013;
39 Mowo et al. 2013).

40 Formal-informal institutional interaction could take different shapes such as: complementary,
41 accommodating, competing, and substitutive. There are many examples when formal institutions
42 might obstruct, change, and hinder informal institutions (Rahman et al. 2014; Helmke and Levitsky
43 2004; Bennett 2013) (Osei-Tutu et al. 2014). Similarly, informal institutions can replace, undermine,
44 and reinforce formal institutions (Grzymala-Busse 2010). In the absence of formal institutions,
45 informal institutions gain importance requiring focus in relation to natural resources management and

1 rights protection (Estrin and Prevezer 2011; Helmke and Levitsky 2004; Kangalawe.R.Y.M, Noe.C,
2 Tungaraza.F.S.K 2014; Sauerwald and Peng 2013; Zoogah et al. 2015).

3 Community forestry comprises 22% of forests in tropical countries in contrast to large-scale industrial
4 forestry (Hajjar et al. 2013) and is managed with informal institutions ensuring a sustainable flow of
5 forest products and income utilising traditional ecological knowledge to determine access to resources
6 (Singh et al. 2018). Policies that create an open platform for local debates and allow actors their own
7 active formulation of rules strengthen informal institutions. Case studies in Zambia, Mali, Indonesia
8 and Bolivia confirm that enabling factors for advancing the local ownership of resources and crafting
9 durability of informal rules require recognition in laws, regulations and policies of the state (Haller et
10 al. 2016).

11 **7.5.2. Decision Making, Timing, Risk, and Uncertainty**

12 This section assesses decision making literature concluding advances in methods have been made in
13 the face of conceptual risk literature and together with a synthesis of empirical evidence, near term
14 decisions have significant impact on costs.

15 **7.5.2.1. Problem Structuring**

16 Structured decision making occurs when there is scientific knowledge about cause and effect, little
17 uncertainty, and agreement exists on values and norms relating to an issue (Hurlbert and Gupta 2016).
18 This decision space is situated within the “known” space where cause and effect is understood and
19 predictable (although uncertainty is not quite zero) (French 2015). Figure 7.5 displays the structured
20 problem area in the bottom left corner corresponding with the ‘known’ decision making space.
21 Decision making surrounding quantified risk assessment and risk management (7.4.3.1) occurs within
22 this decision making space. Examples in the land and climate area include cost benefit analysis
23 surrounding implementation of irrigation projects (Batie 2008) or adopting soil erosion practices by
24 agricultural producers based on anticipated profit (Hurlbert 2018b). Comprehensive risk management
25 also occupies this decision space (Papathoma-Köhle et al. 2016), encompassing risk assessment,
26 reduction, transfer, retention, emergency preparedness and response, and disaster recovery by
27 combining quantified proactive and reactive approaches (Fra.Paleo 2015) (see 7.4.3).

28 A moderately structured decision space is characterised as one where there is either some
29 disagreement on norms, principles, ends and goals in defining a future state or there is some
30 uncertainty surrounding land and climate including land use, observations of land use changes, early
31 warning and decision support systems, model structures, parameterisations, inputs, or from unknown
32 futures informing integrated assessment models and scenarios (see Chapter 1, 1.2.2 and Cross chapter
33 Box 1 on Scenarios). Environmental decision making often takes place in this space where there is
34 limited information and ability to process it, and individual stakeholders make different decisions on
35 the best future course of action (Waas et al. 2014) (*medium confidence*) (Hurlbert and Gupta 2016,
36 2015; Hurlbert 2018b). Figure 7.5 displays the moderately structured problem space characterised by
37 disagreement surrounding norms on the top left hand side. This corresponds with the complex
38 decision making space, the realm of social sciences and qualitative knowledge, where cause and effect
39 is difficult to relate with any confidence (French 2013).

40 The moderately structured decision space characterised by uncertainty surrounding land and climate
41 on the bottom right hand side of Figure 7.5 as well and corresponds to the knowable decision making
42 space, where the realm of scientific inquiry investigates cause and effects. Here there is sufficient
43 understanding to build models, but not enough understanding to define all parameters (French 2015).

44 The top right hand corner of Figure 7.5 corresponds to the ‘unstructured’ problem or chaotic space
45 where patterns and relationships are difficult to discern and unknown unknowns reside (French 2013).
46 It is in the complex but knowable space, the structured and moderately structured space, that decision
47 making under uncertainty occurs.

7.5.2.2. Decision Making Tools

Decisions can still be made despite uncertainty (*medium confidence*), and a wide range of possible approaches are emerging to support decision-making under uncertainty (Jones et al. 2014), applied both to adaptation and mitigation decisions.

Traditional approaches for economic appraisal, including cost benefit analysis and cost effectiveness analysis referred to in 7.5.2.1 do not handle or address uncertainty well (Hallegatte 2009) (Farber 2015) and favour decisions with short term benefits (see Cross-Chapter Box 10: Economic Dimensions in this chapter). Alternative economic decision making approaches aim to better incorporate uncertainty while still delivering adaptation goals, by selecting projects that meet their purpose across a variety of plausible futures (Hallegatte et al. 2012); so-called ‘robust’ decision-making approaches. These are designed to be less sensitive to uncertainty about the future (Lempert and Schlesinger 2000).

Much of the research for adaptation to climate change has focused around three main economic approaches: Real Options Analysis, Portfolio Analysis, and Robust Decision-Making. Real Options Analysis develops flexible strategies that can be adjusted when additional climate information becomes available. It is most appropriate for large irreversible investment decisions. Applications to climate adaptation are growing quickly, with most studies addressing flood risk and sea-level rise (Gersonius et al. 2013; Woodward et al. 2014; Dan 2016), but studies in land use decisions are also emerging, including identifying the optimal time to switch land use in a changing climate (Sanderson et al. 2016) and water storage (Sturm et al. 2017; Kim et al. 2017). Portfolio analysis aims to reduce risk by diversification, by planting multiple species rather than only one, in forestry (Knoke et al. 2017) or crops (Ben-Ari and Makowski 2016), for example, or in multiple locations. There may be a trade-off between robustness to variability and optimality (Yousefpour and Hanewinkel 2016; Ben-Ari and Makowski 2016); but this type of analysis can help identify and quantify trade-offs. Robust Decision Making identifies how different strategies perform under many climate outcomes, also potentially trading off optimality for resilience (Lempert 2013).

Multi-criteria decision making continues to be an important tool in the land-use sector, with the capacity to simultaneously consider multiple goals across different domains (e.g., economic, environmental, social) (Bausch et al. 2014; Alrø et al. 2016), and is thus useful as a mitigation as well as an adaptation tool. Life-cycle assessment (LCA) can also be used to evaluate emissions across a system (for example in livestock production (McClelland et al. 2018)) and identify areas to prioritise for reductions. Bottom-up Marginal Abatement Cost Curves calculate the most cost-effective cumulative potential for mitigation across different options (Eory et al. 2018).

In the climate adaptation literature, these tools may be used in adaptive management (see 7.5.4), using a monitoring, research, evaluation and learning process (cycle) to improve future management strategies (Tompkins and Adger 2004). More recently these techniques have been advanced with iterative risk management (IPCC 2014a) (see 7.4.1, 7.4.7), adaptation pathways (Downing 2012), and dynamic adaptation pathways (Haasnoot et al. 2013) (see 7.6.3). Decision making tools can be selected and adapted to fit the specific land and climate problem and decision making space. For instance, dynamic adaptation pathways processes (Haasnoot et al. 2013; Wise et al. 2014) identify and sequence potential actions based on alternative potential futures and are situated within the complex, unstructured space (see Figure 7.5). Decisions are made based on trigger points, linked to indicators and scenarios, or changing performance over time (Kwakkel et al. 2016). A key characteristic of these pathways is rather than making irreversible decisions now, decisions evolve over time, accounting for learning (see 7.6.4), knowledge, and values. Combining Dynamic Adaptive Pathways and a form of Real Options Analysis with Multi Criteria Decision Analysis has enabled changing risk over time to be included in assessment of adaptation options through a participatory learning process in New Zealand (Lawrence et al. 2019).

1 Scenario analysis is also situated within the complex, unstructured space (although unlike adaptation
2 pathways, it does not allow for changes in pathway over time) and is important for identifying
3 technology and policy instruments to ensure spatial-temporal coherence of land use allocation
4 simulations with scenario storylines (Brown and Castellazzi 2014) and identifying technology and
5 policy instruments for mitigation of land degradation (Fleskens et al. 2014).

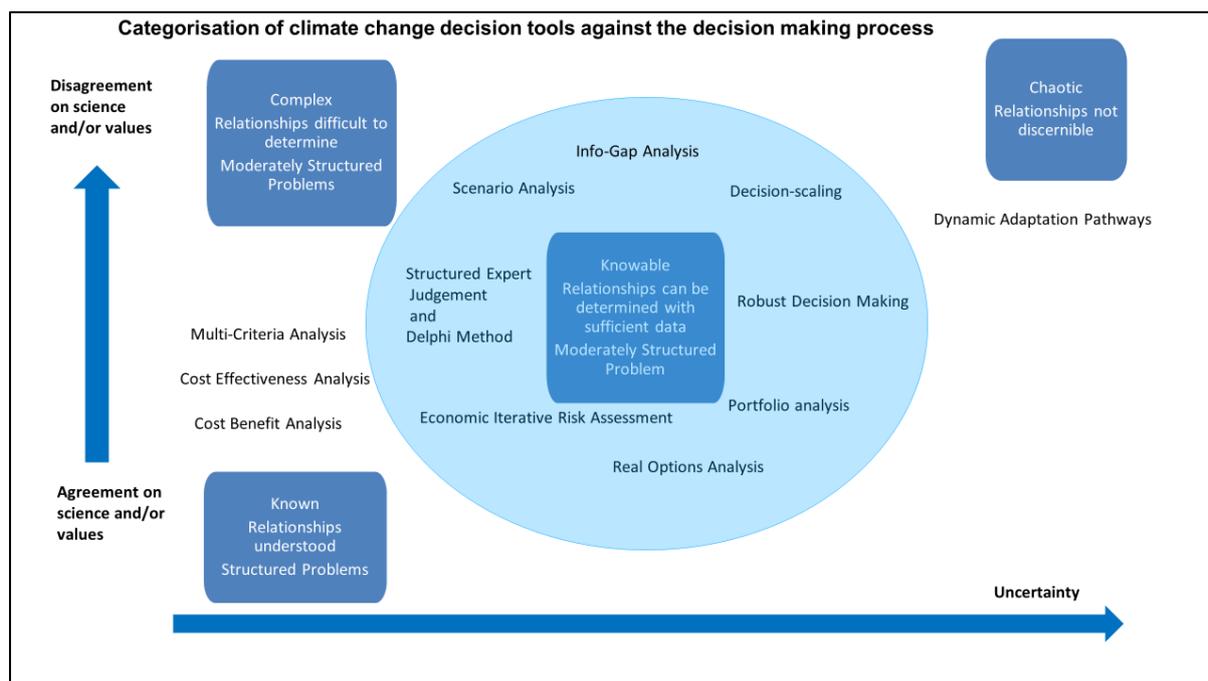
6 While economics is usually based on the idea of a self-interested, rational agent, more recently
7 insights from psychology are being used to understand and explain human behaviour in the field of
8 behavioural economics (Shogren and Taylor 2008; Kesternich et al. 2017), illustrating how a range of
9 cognitive factors and biases can affect choices (Valatin et al. 2016). These insights can be critical in
10 supporting decision-making that will lead to more desirable outcomes relating to land and climate
11 change. One example of this is ‘policy nudges’ (Thaler and Sunstein 2008) which can ‘shift choices
12 in socially desirable directions’ (Valatin et al. 2016). Tools can include framing tools, binding pre-
13 commitments, default settings, channel factors, or broad choice bracketing (Wilson et al. 2016).
14 Although relatively few empirical examples exist in the land sector, there is evidence that nudges
15 could be applied successfully, for example in woodland creation (Valatin et al. 2016) and agri-
16 environmental schemes (Kuhfuss et al. 2016) (*Medium certainty, low evidence*). Consumers can be
17 ‘nudged’ to consume less meat (Rozin et al. 2011) or to waste food less (Kallbekken and Sælen 2013).

18 Programmes supporting and facilitating desired practices can have success at changing behaviour,
19 particularly if they are co-designed by the end-users (farmers, foresters, land-users) (*medium
20 evidence, high agreement*). Programmes that focus on demonstration or trials of different adaptation
21 and mitigation measures, and facilitate interaction between farmers, industry specialists are perceived
22 as being successful (Wreford et al. 2017; Hurlbert 2015b) but systematic evaluations of their success
23 at changing behaviour are limited (Knook et al. 2018).

24 Different approaches to decision making are appropriate in different contexts. Dittrich et al. (2017)
25 provide a guide to the appropriate application in different contexts for adaptation in the livestock
26 sector in developed countries. While considerable advances have been made in the theoretical
27 approaches, a number of challenges arise when applying these in practice, and partly relate to the
28 necessity of assigning probabilities to climate projects, and the complexity of the approaches being a
29 prohibitive factor beyond academic exercises. Formalised expert judgement can improve how
30 uncertainty is characterised (Kunreuther et al. 2014) and these methods have been improved utilising
31 Bayesian belief networks to synthesise expert judgements and include fault trees and reliability block
32 diagrams to overcome standard reliability techniques (Sigurdsson et al. 2001) as well as mechanisms
33 incorporating transparency (Ashcroft et al. 2016).

34 It may also be beneficial to combine decision making approaches with the precautionary principle, or
35 the idea that lack of scientific certainty is not to postpone action when faced with serious threats or
36 irreversible damage to the environment (Farber 2015). The precautionary principle requires cost
37 effective measures to address serious but uncertain risks (Farber 2015). It supports a rights based
38 policy instrument choice as consideration is whether actions or inactions harm others moving beyond
39 traditional risk management policy considerations that surround net benefits (Etkin et al. 2012).
40 Farber, (2015) concludes the principle has been successfully applied in relation to endangered species
41 and situations where climate change is a serious enough problem to justify some response. There is
42 *medium confidence* that combining the precautionary principle with integrated assessment models,
43 risk management, and cost benefit analysis in an integrated, holistic manner, together would be a good
44 combination of decision making tools supporting sustainable development (Farber 2015; Etkin et al.
45 2012).

1



2

3

4

Figure 7.5 Structural and Uncertain Decision Making

7.5.2.3. Cost and timing of action

The Cross-Chapter Box 10 on Economics Dimensions deals with the costs and timing of action. In terms of policies, not only is timing important, but the type of intervention itself can influence returns (*high evidence, high agreement*). Policy packages that make people more resilient - expanding financial inclusion, disaster risk and health insurance, social protection and adaptive safety nets, contingent finance and reserve funds, and universal access to early warning systems (see 7.4.1, 7.6.3) – could save USD 100 billion a year, if implemented globally (Hallegatte et al. 2017). In Ethiopia, Kenya and Somalia, every 1 USD spent on safety net/resilience programming results in net benefits of between USD 2.3 and 3.3 (Venton 2018). Investing in resilience building activities, which increase household income by USD 365 to 450 per year in these countries, is more cost effective than providing ongoing humanitarian assistance.

There is a need to further examine returns on investment for land-based adaptation measures, both in the short and long term. Other outstanding questions include identifying specific triggers for early response. Food insecurity, for example, can occur due to a mixture of market and environmental factors (changes in food prices, animal or crop prices, rainfall patterns) (Venton 2018). The efficacy of different triggers, intervention times and modes of funding are currently being evaluated (see for example forecast based finance study (Alverson and Zommers 2018)). To reduce losses and maximise returns on investments, this information can be used to develop: 1) coordinated, agreed plans for action; 2) a clear, evidence-based decision-making process, and; 3) financing models to ensure that the plans for early action can be implemented (Clarke and Dercon 2016a).

25

7.5.3. Best practices of decision making toward sustainable land management

Sustainable land management is a strategy and also an outcome (Waas et al. 2014) and decision making practices are fundamental in achieving it as an outcome (*medium evidence, medium agreement*). Sustainable land management decision making is improved (*medium evidence and high*

29

1 *agreement*) with ecological service mapping with three characteristics: robustness (robust modelling,
2 measurement, and stakeholder-based methods for quantification of ecosystem service supply, demand
3 and/or flow, as well as measures of uncertainty and heterogeneity across spatial and temporal scales
4 and resolution); transparency (to contribute to clear information-sharing and the creation of linkages
5 with decision support processes); and relevancy to stakeholders (people-central in which stakeholders
6 are engaged at different stages) (Willemsen et al. 2015; Ashcroft et al. 2016). Practices that advance
7 sustainable land management include remediation practices as well as critical interventions that are
8 reshaping norms and standards, joint implementation, experimentation, and integration of rural actors'
9 agency in analysis and approaches in decision-making (Hou and Al-Tabbaa 2014). Best practices are
10 identified in the literature after their implementation demonstrates effectiveness at improving water
11 quality, the environment, or reducing pollution (Rudolph et al. 2015; Lam et al. 2011).

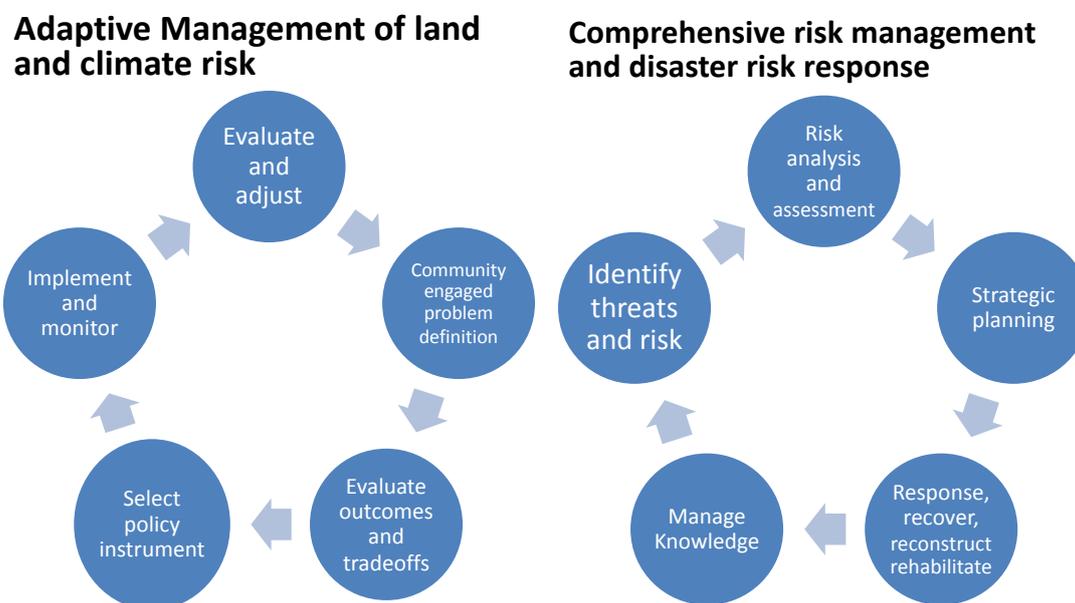
12 There is *medium evidence and medium agreement* about what factors consistently determine the
13 adoption of agricultural best management practices (Herendeen and Glazier 2009) and these
14 positively correlate to education levels, income, farm size, capital, diversity, access to information,
15 and social networks. Attending workshops for information and trust in crop consultants are also
16 important factors in adoption of best management practices (Ulrich-Schad, J.D., Garcia de Jalon, S.,
17 Babin, N., Paper, A. 2017; Baumgart-Getz et al. 2012). More research is needed on the sustained
18 adoption of these factors over time (Prokopy et al. 2008).

19 There is *medium evidence and high agreement* that sustainable land management practices and
20 incentives require mainstreaming into relevant policy; appropriate market based approaches, including
21 payment for ecosystem services and public private partnerships, need better integration into payment
22 schemes (Tengberg et al. 2016). There is *medium evidence and high agreement* that many of the best
23 sustainable land management decisions are made with the participation of stakeholders and social
24 learning (Section 7.6.4) (Stringer and Dougill 2013). As stakeholders may not be in agreement, either
25 practices of mediating agreement, or modelling that depicts and mediates the effects of stakeholder
26 perceptions in decision making may be applicable (Hou 2016; Wiggering and Steinhardt 2015).

27 **7.5.4. Adaptive management**

28 Adaptive management is an evolving approach to natural resource management founded on decision
29 making approaches in other fields (such as business, experimental science, and industrial ecology)
30 (Allen et al. 2011; Williams 2011) and decision making that overcomes management paralysis and
31 mediates multiple stakeholder interests through use of simple steps. (Adaptive governance considers a
32 broader socio-ecological system that includes the social context that facilitates adaptive management
33 (Chaffin et al. 2014)). Adaptive management steps include evaluating a problem and integrating
34 planning, analysis and management into a transparent process to build a road map focused on
35 achieving fundamental objectives. Requirements of success are clearly articulated objectives, the
36 explicit acknowledgment of uncertainty, and a transparent response to all stakeholder interests in the
37 decision making process (Allen et al. 2011). Adaptive management builds on this foundation by
38 incorporating a formal iterative process acknowledging uncertainty and achieving management
39 objectives through a structured feedback process that includes stakeholder participation (see 7.6.4)
40 (Foxon et al. 2009). In the adaptive management process the problem and desired goals are identified,
41 evaluation criteria formulated, the system boundaries and context are ascertained, tradeoffs evaluated,
42 decisions are made regarding responses and policy instruments, which are implemented, and
43 monitored, evaluated and adjusted (Allen et al. 2011). The implementation of policy strategies and
44 monitoring of results occurs in a continuous management cycle of monitoring, assessment and
45 revision (Hurlbert 2015b; Newig et al. 2010; Pahl-Wostl et al. 2007) as illustrated in Figure 7.6.

Adaptive Risk Governance



1

2

Figure 7.6 Adaptive Governance, Management, and Comprehensive Iterative Risk Management.

3

Source: Adapted from (Ammann 2013; Allen et al. 2011)

4 A key focus on adaptive management is the identification and reduction of uncertainty (as described
 5 in Chapter 1, 1.2.2 and Cross-Chapter Box 1 on Scenarios) and partial controllability whereby policies
 6 used to implement an action are only indirectly responsible (for example setting a harvest rate)
 7 (Williams 2011). There is *medium evidence and high agreement* that adaptive management is an ideal
 8 method to resolve uncertainty when uncertainty and controllability (resources will respond to
 9 management) are both high (Allen et al. 2011). Where uncertainty is high, but controllability is low,
 10 developing and analysing scenarios may be more appropriate (Allen et al. 2011). Anticipatory
 11 governance has developed combining scenarios and forecasting in order to creatively design strategy
 12 to address complex, fuzzy and wicked challenges (Ramos 2014; Quay 2010) (see 7.5). Even where
 13 there is low controllability, such as in the case of climate change, adaptive management can help
 14 mitigate impacts including changes in water availability and shifting distributions of plants and
 15 animals (Allen et al. 2011).

16 There is *medium evidence and high agreement* that adaptive management can help reduce
 17 anthropogenic impacts of changes of land and climate including: species decline and habitat loss
 18 (participative identification, monitoring, and review of species at risk as well as decision making
 19 surrounding protective measures) (Fontaine 2011; Smith 2011) including quantity and timing of
 20 harvest of animals (Johnson 2011a), human participation in natural resource-based recreational
 21 activities including selection fish harvest quotas and fishing seasons from year to year (Martin and
 22 Pope 2011), managing competing interests of land use planners and conservationists in public lands
 23 (Moore et al. 2011), managing endangered species and minimising fire risk through land cover
 24 management (Breininger et al. 2014), land use change in hardwood forestry through mediation of
 25 hardwood plantation forestry companies and other stakeholders including those interested in water,

1 environment or farming (Leys and Vanclay 2011), and sustainable land management protecting
2 biodiversity, increasing carbon storage, and improving livelihoods (Cowie et al. 2011). There is
3 *medium evidence and medium agreement* that despite abundant literature and theoretical explanation,
4 there has remained imperfect realisation of adaptive management because of several challenges: lack
5 of clarity in definition and approach, few success stories on which to build an experiential base
6 practitioner knowledge of adaptive management, paradigms surrounding management, policy and
7 funding that favour reactive approaches instead of the proactive adaptive management approach,
8 shifting objectives that do not allow for the application of the approach, and failure to acknowledge
9 social uncertainty (Allen et al. 2011). Adaptive management includes participation (7.6.4), the use of
10 indicators (7.5.5), in order to avoid maladaptation and trade-offs while maximising synergies (7.5.6).

11 **7.5.5. Performance indicators**

12 Measuring performance is important in adaptive management decision-making, policy instrument
13 implementation, and governance and can help evaluate policy effectiveness (*medium evidence, high*
14 *agreement*) (Wheaton and Kulshreshtha 2017; Bennett and Dearden 2014; Oliveira Júnior et al. 2016;
15 Kaufmann 2009). Indicators can relate to specific policy problems (climate mitigation, land
16 degradation), sectors (agriculture, transportation etc.), and policy goals (SDGs, food security).

17 It is necessary to monitor and evaluate the effectiveness and efficiency of performing climate actions
18 to ensure the long-term success of *climate* initiatives or plans. Measurable indicators are useful for
19 climate policy development and decision-making process since they can provide quantifiable
20 information regarding the progress of climate actions. The Paris Agreement (UNFCCC 2015)
21 focused on reporting the progress of implementing countries' pledges, i.e., NDCs and national
22 adaptation needs in order to examine the aggregated results of mitigation actions that have already
23 been implemented. For the case of measuring progress toward achieving land degradation neutrality,
24 it was suggested to use land-based indicators, i.e., trend in land cover, trends in land productivity or
25 functioning of the land, and trends in carbon stock above and below ground (Cowie et al. 2018a).
26 There is *medium evidence and high agreement* that indicators for measuring biodiversity and
27 ecosystem services in response to governance at local to international scale meet the criteria of
28 parsimony and scale specificity, are linked to some broad social, scientific and political consensus on
29 desirable states of ecosystems and biodiversity, and include normative aspects such as environmental
30 justice or socially just conservation (Layke 2009) (Van Oudenhoven et al. 2012) (Turnhout et al.
31 2014)(Häyhä and Franzese 2014), (Guerry et al. 2015)(Díaz et al. 2015).

32 Important in making choices of metrics and indicators is understanding that the science, linkages and
33 dynamics in systems are complex, not amenable to be addressed by simple economic instruments, and
34 are often unrelated to short-term management or governance scales (Naeem et al. 2015) (Muradian
35 and Rival 2012). Thus, ideally stakeholders participate in the selection and use of indicators for
36 biodiversity and ecosystem services and monitoring impacts of governance and management regimes
37 on land-climate interfaces. The adoption of non-economic approaches that are part of the emerging
38 concept of Nature's Contributions to People (NCP) could potentially elicit support for conservation
39 from diverse sections of civil society (Pascual et al. 2017).

40 Recent studies increasingly incorporate the role of stakeholders and decision makers in selection of
41 indicators for land systems (Verburg et al. 2015) including sustainable agriculture (Kanter et al.
42 2016), bioenergy sustainability (Dale et al. 2015), desertification (Liniger et al. 2019), and
43 vulnerability (Debortoli et al. 2018). Kanter et al. (2016) propose a four-step cradle-to-grave approach
44 for agriculture trade-off analysis, which involves co-evaluation of indicators and trade-offs with both
45 stakeholders and decision-makers.

46

1 **7.5.6. Maximising Synergies and Minimising Trade-offs**

2 Synergies and trade-offs to address land and climate related measures are identified and discussed in
3 Chapter 6. Here we outline policies supporting Chapter 6 response options (see Table 7.5), and
4 discuss synergies and trade-offs in policy choices and interactions among policies. Trade-offs will
5 exist between broad policy approaches. For example, while legislative and regulatory approaches may
6 be effective at achieving environmental goals, they may be costly and ideologically unattractive in
7 some countries. Market-driven approaches such as carbon pricing are cost effective ways to reduce
8 emissions, but may not be favoured politically and economically (see 7.4.4). Information provision
9 involves little political risk or ideological constraints, but behavioural barriers may limit their
10 effectiveness (Henstra 2016). This level of trade-off is often determined by the prevailing political
11 system.

12 Synergies and trade-offs also result from interaction between policies (policy interplay (Urwin and
13 Jordan 2008)) at different levels of policy (vertical) and across different policies (horizontal) (see also
14 section on policy coherence, 7.4.8)). If policy mixes are designed appropriately, acknowledging and
15 incorporating trade-offs and synergies, they are better placed to deliver an outcome such as
16 transitioning to sustainability (Howlett and Rayner 2013; Huttunen et al. 2014) (*medium evidence and*
17 *medium agreement*). However, there is *limited evidence and medium agreement* that evaluating
18 policies for coherence in responding to climate change and its impacts is not occurring, and policies
19 are instead reviewed in a fragmented manner (Hurlbert and Gupta 2016).

20

1

Table 7.5 Selection of Policies/Programmes/Instruments that support response options

| Category | Integrated Response Option | Policy instrument supporting response option |
|--|---|--|
| Land management in agriculture | Increased food productivity | Investment in agricultural research for crop and livestock improvement, agricultural technology transfer, inland capture fisheries and aquaculture {7.4.7} agricultural policy reform and trade liberalisation |
| | Improved cropland, grazing, and livestock management | Environmental farm programs/agri-environment schemes, water efficiency requirements and water transfer {3.7.5}, extension services |
| | Agroforestry | Payment for ecosystem services {7.4.6} |
| | Agricultural diversification | Elimination of agriculture subsidies {5.7.1}, environmental farm programs, agri-environmental payments {7.4.6}, rural development programmes |
| | Reduced grassland conversion to cropland | Elimination of agriculture subsidies, remove insurance incentives, ecological restoration {7.4.6} |
| | Integrated water management | Integrated governance {7.6.2}, multi-level instruments {7.4.1} |
| Land management in forests | Forest management, Reduced deforestation and degradation, Reforestation and forest restoration, Afforestation | REDD+, forest conservation regulations, payments for ecosystem services, recognition of forest rights and land tenure {7.4.6}, adaptive management of forests {7.5.4}, land use moratoriums, reforestation programs and investment {4.9.1} |
| Land management of soils | Increased soil organic carbon content, Reduced soil erosion, Reduced soil salinisation, Reduced soil compaction, Biochar addition to soil | Land degradation neutrality {7.4.5}, drought plans, flood plans, flood zone mapping {7.4.3}, technology transfer {7.4.4}, land use zoning {7.4.6}, ecological service mapping and stakeholder based quantification {7.5.3}, environmental farm programs/agri-environment schemes, water efficiency requirements and water transfer {3.7.5} |
| Land management in all other ecosystems | Fire management | Fire suppression, prescribed fire management, mechanical treatments {7.4.3} |
| | Reduced landslides and natural hazards | Land use zoning {7.4.6} |
| | Reduced pollution - acidification | Environmental regulations, Climate mitigation (carbon pricing) {7.4.4} |
| | Management of invasive species / encroachment | Invasive species regulations, trade regulations {5.7.2, 7.4.6} |
| | Restoration and reduced conversion of coastal wetlands | Flood zone mapping {7.4.3}, land use zoning {7.4.6} |
| | Restoration and reduced conversion of peatlands | Payment for ecosystem services {7.4.6; 7.5.3}, standards and certification programs {7.4.6}, land use moratoriums |
| CDR Land management | Biodiversity conservation | Conservation regulations, protected areas policies |
| | Enhanced weathering of minerals | No data |
| Demand management | Bioenergy and BECCS | Standards and certification for sustainability of biomass and land use {7.4.6} |
| | Dietary change | Awareness campaigns/education, changing food choices through nudges, synergies with health insurance and policy {5.7.2} |
| Supply management | Reduced post-harvest losses | Agricultural business risk programs {7.4.8}; regulations to reduce and taxes on food waste, Improved shelf life, circularising the economy to produce substitute goods, carbon pricing, sugar/fat taxes {5.7.2} |
| | Reduced food waste (consumer or retailer), Material substitution | |
| | Sustainable sourcing | Food labelling, innovation to switch to food with lower environmental footprint, public procurement policies {5.7.2}, standards and certification programs {7.4.6} |
| | Management of supply chains | Liberalised international trade {5.7.2}, food purchasing and storage policies of governments, standards and certification programs {7.4.6}, regulations on speculation in food systems |
| | Enhanced urban food systems | Buy local policies; land use zoning to encourage urban agriculture, nature-based solutions and green infrastructure in cities; incentives for technologies like vertical farming |
| Risk management | Improved food processing and retailing, Improved energy use in food systems | Agriculture emission trading {7.4.4}; investment in research and development for new technologies; certification |
| | Management of urban sprawl | Land use zoning {7.4.6} |
| | Livelihood diversification | Climate-smart agriculture policies, adaptation policies, extension services {7.5.6} |
| | Disaster risk management Risk sharing instruments | Disaster risk reduction {7.5.4; 7.4.3}, adaptation planning Insurance, iterative risk management, Cat bonds, risk layering, contingency funds {7.4.3}, agriculture business risk portfolios {7.4.8} |

2

Cross-Chapter Box 9 on Illustrative Climate and Land Pathways

Katherine Calvin (The United States of America), Edouard Davin (France/Switzerland), Margot Hurlbert (Canada), Jagdish Krishnaswamy (India), Alexander Popp (Germany), Prajal Pradhan (Nepal/Germany)

Future development of socioeconomic factors and policies influence the evolution of the land-climate system, among others in terms of the land used for agriculture and forestry. Climate mitigation policies can also have a major impact on land use, especially in scenarios consistent with the climate targets of the Paris Agreement. This includes the use of bio-energy or Carbon Dioxide Removal (CDR), such as bioenergy with carbon dioxide capture and storage (BECCS) and afforestation. Land-based mitigation options have implications for GHG fluxes, desertification, land degradation, food insecurity, ecosystem services and other aspects of sustainable development.

Illustrative Futures

The three illustrative futures are based on the Shared Socioeconomic Pathways (SSPs; (O'Neill et al. 2014c; Riahi et al. 2017b; Popp et al. 2017; Rogelj et al. 2018b); Cross-Chapter Box 1 in Chapter 1). SSP1 is a scenario with a broad focus on sustainability including a focus on human development, technological development, nature conservation, globalised economy, economic convergence and early international cooperation including moderate levels of trade. The scenario assumes a low population growth, relatively high agricultural yields and a move towards less-meat intensive diets (van Vuuren et al. 2017b). Dietary change and reductions in food waste reduce agricultural demands and well-managed land systems enable reforestation and/or afforestation. SSP2 is a scenario in which societal as well as technological development follows historical patterns (Fricko et al. 2017). Land-based CDR is achieved through bioenergy and BECCS, and to a lesser degree by afforestation and reforestation. SSP3 is a scenario with limited technological progress and land-use regulation. Agricultural demands are high due to resource-intensive consumption and a regionalised world leads to reduced flows for agricultural goods. In SSP3, forest mitigation activities and abatement of agricultural GHG emissions are limited due to major implementation barriers such as low institutional capacities in developing countries and delayed as a consequence of low international cooperation (Fujimori et al. 2017a). Emissions reductions are achieved primarily through the energy sector, including the use of bioenergy and BECCS.

Policies in the Illustrative Futures

SSPs are complemented by a set of shared policy assumptions (Kriegler et al. 2014), indicating the types of policies that may be implemented in each future world. IAMs represent the effect of these policies on the economy, energy system, land use and climate with the caveat that they are assumed to be effective or in some cases the policy goals (e.g., dietary change) are imposed rather than explicitly modelled. In the real world, there are various barriers that can make policy implementation more difficult (see 7.4.9). These barriers will be generally higher in SSP3 than SSP1.

SSP1: A number of policies could support this SSP1 future including: effective carbon pricing, emission trading schemes (including net CO₂ emissions from agriculture), carbon taxes, regulations limiting GHG emissions and air pollution, forest conservation (mix of land-sharing and land sparing) through participation, incentives for ecosystem services and secure tenure, and protecting the environment, microfinance, crop and livelihood insurance, agriculture extension services, agricultural production subsidies, low export tax and import tariff rates on agricultural goods, dietary awareness campaigns, regulations to reduce and taxes on food waste, improved shelf life, sugar/fat taxes, and instruments supporting sustainable land management including payment for ecosystem services, land use zoning, REDD+, standards and certification for sustainable biomass production practices, legal reforms on land ownership and access, legal aid, legal education, including reframing these policies as

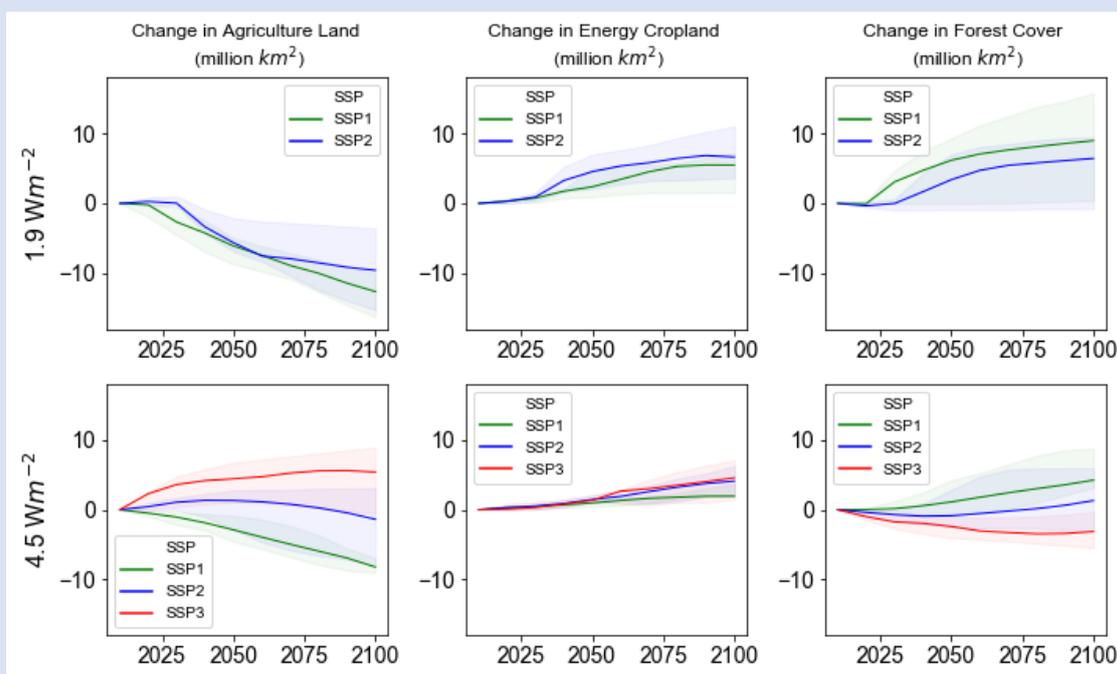
entitlements for women and small agricultural producers (rather than sustainability) (O'Neill et al. 2017; van Vuuren et al. 2017b) (see 7.4).

SSP2: The same policies that support the SSP1 could support the SSP2 but may be less effective and only moderately successful. Policies may be challenged by adaptation limits (7.4.9), inconsistency in formal and informal institutions in decision making (7.5.1) or result in maladaptation (7.4.7). Moderately successful sustainable land management policies result in some land competition. Land degradation neutrality is moderately successful. Successful policies include those supporting bioenergy and BECCS (Rao et al. 2017; Riahi et al. 2017b; Fricko et al. 2017) (see 7.4.6).

SSP3: Policies that exist in SSP1 may or may not exist in SSP3, and are ineffective (O'Neill et al. 2014c). There are challenges to implementing these policies, as in SSP2. In addition, ineffective sustainable land management policies result in competition for land between agriculture and mitigation. Land degradation neutrality is not achieved (Riahi et al. 2017b). Successful policies include those supporting bioenergy and BECCS (see 7.4.6) (Kriegler et al. 2017; Fujimori et al. 2017a; Rao et al. 2017). Demand side food policies are absent and supply side policies predominate. There is no success in advancing land ownership and access policies for agricultural producer livelihood (7.6.5).

Land use and land cover change

Agricultural area in SSP1 declines as a result of the low population growth, agricultural intensification, low meat consumption, and low food waste. In contrast, SSP3 has high population and strongly declining rates of crop yield growth over time, resulting in increased agricultural land area. The SSP2 falls somewhere in between, with its modest growth in all factors. In the climate policy scenarios consistent with the Paris Agreement, bioenergy/BECCS and reforestation/afforestation play an important role in SSP1 and SSP2. The use of these options, and the impact on land, is larger in scenarios that limit radiative forcing in 2100 to 1.9 Wm^{-2} than in the 4.5 Wm^{-2} scenarios. In SSP3, the expansion of land for agricultural production implies that the use of land-related mitigation options is very limited, and the scenario is characterised by continued deforestation.



Cross-Chapter Box 9 Figure 1: Changes in agricultural land (left), energy cropland (middle) and forest cover (right) under three different SSPs (colours) and two different warming levels (rows). Agricultural land includes both pasture and non-energy cropland. Colours indicate SSPs, with SSP1 shown in green,

SSP2 in blue, and SSP3 in red. Shaded area show the range across all IAMs; lines show the median across all models. Models are only included in a figure if they provided results for all SSPs in that panel. There is no SSP3 in the top row, as 1.9 Wm^{-2} is infeasible in this world. Data is from an update of the IAMC Scenario Explorer developed for the SR15 (Huppmann et al. 2018; Rogelj et al. 2018a).

Implications for mitigation and other land challenges

The combination of baseline emissions development, technology options, and policy support makes it much easier to reach the climate targets in the SSP1 scenario than in the SSP3 scenario. As a result, carbon prices are much higher in SSP3 than in SSP1. In fact, the 1.9 Wm^{-2} target was found to be infeasible in the SSP3 world (Cross-Chapter Box 9 Table 1). Energy system CO_2 emissions reductions are greater in the SSP3 than in the SSP1 to compensate for the higher land-based CO_2 emissions.

Accounting for mitigation and socioeconomics alone, food prices (an indicator of food insecurity) are higher in SSP3 than in the SSP1 and higher in the 1.9 Wm^{-2} than in the 4.5 Wm^{-2} (Cross-Chapter Box 9 Table 1). Forest cover is higher in the SSP1 than the SSP3 and higher in the 1.9 Wm^{-2} than in the 4.5 Wm^{-2} . Water withdrawals and water scarcity are in general higher in the SSP3 than the SSP1 (Hanasaki et al. 2013a; Graham et al. 2018b) and higher in scenarios with more bioenergy (Hejazi et al. 2014c); however, these indicators have not been quantified for the specific SSP-RCP combinations discussed here.

Climate change, results in higher impacts and risks in the 4.5 Wm^{-2} world than in the 1.9 Wm^{-2} world for a given SSP and these risks are exacerbated in SSP3 compared to SSP1 and SSP2 due to population's higher exposure and vulnerability. For example, the risk of fire is higher in warmer worlds; in the 4.5 Wm^{-2} world, the population living in fire prone regions is higher in the SSP3 (646 million) than in the SSP2 (560 million) (Knorr et al. 2016). Global exposure to multi-sector risk quadruples between the 1.5°C^1 and 3°C and is a factor of six higher in the SSP3- 3°C than in the SSP1- 1.5°C (Byers et al. 2018). Future risks resulting from desertification, land degradation and food insecurity are lower in the SSP1 compared to SSP3 at the same level of warming. For example, the transition moderate to high risk of food insecurity occurs between 1.3 and 1.7°C for the SSP3, but not until 2.5 to 3.5°C in the SSP1 (Section 7.2).

Table 1: Quantitative indicators for the illustrative pathways. Each cell shows the mean, minimum, and maximum value across IAM models for each indicator and each pathway in 2050 and 2100. All IAMs that provided results for a particular pathway are included here. Note that these indicators exclude the implications of climate change. Data is from an update of the IAMC Scenario Explorer developed for the SR15 (Huppmann et al. 2018; Rogelj et al. 2018b).

| | | SSP1 | | SSP2 | | SSP3 | |
|---------------------------------------|------|---|---|---|---|---|---|
| | | 1.9 Wm^{-2} mean (min, max) | 4.5 Wm^{-2} mean (min, max) | 1.9 Wm^{-2} mean (min, max) | 4.5 Wm^{-2} mean (min, max) | 1.9 Wm^{-2} mean (min, max) | 4.5 Wm^{-2} mean (min, max) |
| Population (billion) | 2050 | 8.5 (8.5, 8.5) | 8.5 (8.5, 8.5) | 9.2 (9.2, 9.2) | 9.2 (9.2, 9.2) | N/A | 10.0 (10.0, 10.0) |
| | 2100 | 6.9 (7.0, 6.9) | 6.9 (7.0, 6.9) | 9.0 (9.0, 9.0) | 9.0 (9.1, 9.0) | N/A | 12.7 (12.8, 12.6) |
| Change in GDP per capita (% rel to | 2050 | 170.3 (380.1, | 175.3 (386.2, | 104.3 (223.4, | 110.1 (233.8, | N/A | 55.1 (116.1, 46.7) |

¹ FOOTNOTE: Pathways that limit radiative forcing in 2100 to 1.9 Wm^{-2} result in median warming in 2100 to 1.5°C in 2100 (Rogelj et al. 2018b). Pathways limiting radiative forcing in 2100 to 4.5 Wm^{-2} result in median warming in 2100 above 2.5°C (IPCC 2014).

| | | | | | | | |
|--|------|-------------------------------|-----------------------------|--------------------------------|----------------------------|-----|----------------------------|
| 2010) | | 130.9) | 166.2) | 98.7) | 103.6) | | |
| | 2100 | 528.0 (1358.4, 408.2) | 538.6 (1371.7, 504.7) | 344.4 (827.4, 335.8) | 356.6 (882.2, 323.3) | N/A | 71.2 (159.7, 49.6) |
| Change in forest cover (Mkm ²) | 2050 | 3.4 (9.4, -0.1) | 0.6 (4.2, -0.7) | 3.4 (7.0, -0.9) | -0.9 (2.9, -2.5) | N/A | -2.4 (-1.0, -4.0) |
| | 2100 | 7.5 (15.8, 0.4) | 3.9 (8.8, 0.2) | 6.4 (9.5, -0.8) | -0.5 (5.9, -3.1) | N/A | -3.1 (-0.3, -5.5) |
| Change in cropland (Mkm ²) | 2050 | -1.2 (-0.3, -4.6) | 0.1 (1.5, -3.2) | -1.2 (0.3, -2.0) | 1.2 (2.7, -0.9) | N/A | 2.3 (3.0, 1.2) |
| | 2100 | -5.2 (-1.8, -7.6) | -2.3 (-1.6, -6.4) | -2.9 (0.1, -4.0) | 0.7 (3.1, -2.6) | N/A | 3.4 (4.5, 1.9) |
| Change in energy cropland (Mkm ²) | 2050 | 2.1 (5.0, 0.9) | 0.8 (1.3, 0.5) | 4.5 (7.0, 2.1) | 1.5 (2.1, 0.1) | N/A | 1.3 (2.0, 1.3) |
| | 2100 | 4.3 (7.2, 1.5) | 1.9 (3.7, 1.4) | 6.6 (11.0, 3.6) | 4.1 (6.3, 0.4) | N/A | 4.6 (7.1, 1.5) |
| Change in pasture (Mkm ²) | 2050 | -4.1 (-2.5, -5.6) | -2.4 (-0.9, -3.3) | -4.8 (-0.4, -6.2) | -0.1 (1.6, -2.5) | N/A | 2.1 (3.8, -0.1) |
| | 2100 | -6.5 (-4.8, -12.2) | -4.6 (-2.7, -7.3) | -7.6 (-1.3, -11.7) | -2.8 (1.9, -5.3) | N/A | 2.0 (4.4, -2.5) |
| Change in other natural land (Mkm ²) | 2050 | 0.5 (1.0, -4.9) | 0.5 (1.7, -1.0) | -2.2 (0.6, -7.0) | -2.2 (0.7, -2.2) | N/A | -3.4 (-2.0, -4.4) |
| | 2100 | 0.0 (7.1, -7.3) | 1.8 (6.0, -1.7) | -2.3 (2.7, -9.6) | -3.4 (1.5, -4.7) | N/A | -6.2 (-5.4, -6.8) |
| Carbon price (2010 US\$ per tCO ₂) ^a | 2050 | 510.4 (4304.0, 150.9) | 9.1 (35.2, 1.2) | 756.4 (1079.9, 279.9) | 37.5 (73.4, 13.6) | N/A | 67.2 (75.1, 60.6) |
| | 2100 | 2164.0 (35037.7, 262.7) | 64.9 (286.7, 42.9) | 4353.6 (10149.7, 2993.4) | 172.3 (597.9, 112.1) | N/A | 589.6 (727.2, 320.4) |
| Food price (Index 2010=1) | 2050 | 1.2 (1.8, 0.8) | 0.9 (1.1, 0.7) | 1.6 (2.0, 1.4) | 1.1 (1.2, 1.0) | N/A | 1.2 (1.7, 1.1) |
| | 2100 | 1.9 (7.0, 0.4) | 0.8 (1.2, 0.4) | 6.5 (13.1, 1.8) | 1.1 (2.5, 0.9) | N/A | 1.7 (3.4, 1.3) |
| Increase in Warming above pre-industrial (°C) | 2050 | 1.5 (1.7, 1.5) | 1.9 (2.1, 1.8) | 1.6 (1.7, 1.5) | 2.0 (2.0, 1.9) | N/A | 2.0 (2.1, 2.0) |
| | 2100 | 1.3 (1.3, 1.3) | 2.6 (2.7, 2.4) | 1.3 (1.3, 1.3) | 2.6 (2.7, 2.4) | N/A | 2.6 (2.6, 2.6) |
| Change in per capita demand for food, crops (% rel to 2010) ^b | 2050 | 6.0 (10.0, 4.5) | 9.1 (12.4, 4.5) | 4.6 (6.7, -0.9) | 7.9 (8.0, 5.2) | N/A | 2.4 (5.0, 2.3) |
| | 2100 | 10.1 (19.9, 4.8) | 15.1 (23.9, 4.8) | 11.6 (19.2, -10.8) | 11.7 (19.2, 4.1) | N/A | 2.0 (3.4, -1.0) |
| Change in per capita demand for food, animal products (% rel to 2010) ^{b,c} | 2050 | 6.9 (45.0, -20.5) | 17.9 (45.0, -20.1) | 7.1 (36.0, 1.9) | 10.3 (36.0, -4.2) | N/A | 3.1 (5.9, 1.9) |
| | 2100 | -3.0 (19.8, -27.3) | 21.4 (44.1, -26.9) | 17.0 (39.6, -24.1) | 20.8 (39.6, -5.3) | N/A | -7.4 (-0.7, -7.9) |
| AFOLU CH ₄ Emissions (%) | 2050 | -39.0 (-3.8, -68.9) | -2.9 (22.4, -23.9) | -11.7 (31.4, -59.4) | 7.5 (43.0, -15.5) | N/A | 15.0 (20.1, 3.1) |

| | | | | | | | |
|---|------|-----------------------|-------------------------|------------------------|-------------------------|-----|-------------------------|
| relative to 2010) | 2100 | -60.5 (-41.7, -77.4) | -47.6 (-24.4, -54.1) | -40.3 (33.1, -58.4) | -13.0 (63.7, -45.0) | N/A | 8.0 (37.6, -9.1) |
| AFOLU N ₂ O Emissions (% relative to 2010) | 2050 | -13.1 (-4.1, -26.3) | 0.1 (34.6, -14.5) | 8.8 (38.4, -14.5) | 25.4 (37.4, 5.5) | N/A | 34.0 (50.8, 29.3) |
| | 2100 | -42.0 (4.3, -49.4) | -25.6 (-3.4, -51.2) | -1.7 (46.8, -37.8) | 19.5 (66.7, -21.4) | N/A | 53.9 (65.8, 30.8) |
| Cumulative Energy CO ₂ Emissions until 2100 (GtCO ₂) | | 428.2 (1009.9, 307.6) | 2787.6 (3213.3, 2594.0) | 380.8 (552.8, -9.4) | 2642.3 (2928.3, 2515.8) | N/A | 2294.5 (2447.4, 2084.6) |
| Cumulative AFOLU CO ₂ Emissions until 2100 (GtCO ₂) | | -127.3 (5.9, -683.0) | -54.9 (52.1, -545.2) | -126.8 (153.0, -400.7) | 40.8 (277.0, -372.9) | N/A | 188.8 (426.6, 77.9) |

^a The SSP2-19 is infeasible in two models. One of these models sets the maximum carbon price in the SSP1-19; the carbon price range is smaller for the SSP2-19 as this model is excluded there. Carbon prices are higher in the SSP2-19 than the SSP1-19 for every model that provided both simulations.

^a Food demand estimates include waste.

^b Animal product demand includes meat and dairy.

Summary

Future pathways for climate and land use include portfolios of response and policy options. Depending on the response options included, policy portfolios implemented, and other underlying socioeconomic drivers, these pathways result in different land-use consequences and their contribution to climate change mitigation. Agricultural area declines by more than 5 Mkm² in one SSP but increases by as much as 5 Mkm² in another. The amount of energy cropland ranges from nearly zero to 11 Mkm², depending on the SSP and the warming target. Forest area declines in the SSP3 but increases substantially in the SSP1. Subsequently, these pathways have different implications for risks related to desertification, land degradation, food insecurity, and terrestrial greenhouse gas fluxes, as well as ecosystem services, biodiversity, and other aspects of sustainable development.

7.5.6.1. Trade-offs and Synergies between ES

Unplanned or unintentional trade-offs and synergies between policy driven response options related to ecosystem service (ES) can happen over space (e.g., upstream-downstream, IWM 3.7.5.2) or intensify over time (reduced water in future dry-season due to growing tree plantations, 6.4.1). Trade-offs can occur between two or more ecosystem services (land for climate mitigation vs food 6.2, 6.3, 6.4, Cross-Chapter Box 8: Ecosystem services, Chapter 6; Cross-Chapter Box 9: Illustrative climate and land pathways, Chapter 6), and between scales such as forest biomass based livelihoods versus global ES carbon storage (Chhatre and Agrawal 2009)(*medium evidence, medium agreement*). Tradeoffs can be reversible or irreversible (Rodríguez et al. 2006; Elmqvist et al. 2013)(for example a soil carbon sink is reversible (6.4.1.1)

Although there is *robust evidence* and *high agreement* that ES are important for human well-being, the relationship between poverty alleviation and ES can be surprisingly complex, understudied and dependent on the political economic context; current evidence is largely about provisioning services and often ignores multiple dimensions of poverty (Suich et al. 2015; Vira et al. 2012). Spatially explicit mapping and quantification of stake-holder choices vis-à-vis distribution of various ES can help enhance synergies and reduce trade-offs (Turkelboom et al. 2018; Locatelli et al. 2014)(see 7.5.5).

7.5.6.2. Sustainable Development Goals (SDGs): Synergies and Trade-offs

The SDGs, an international persuasive policy instrument, apply to all countries, and measure sustainable and socially just development of human societies at all scales of governance (Griggs et al. 2013). The UN SDGs rest on the premise that the goals are mutually reinforcing and there exist inherent linkages, synergies and trade-offs (to a greater or lesser extent) between and within the sub-goals (Fuso Nerini et al. 2018; Nilsson et al. 2016b)(Le Blanc 2015). There is *high confidence* that opportunities, trade-offs and co-benefits are context and region specific and depend on a variety of political, national and socio-economic factors (Nilsson et al. 2016b) depending on perceived importance by decision and policy makers (Figure 7.7, Table 7.6 below). Aggregation of targets and indicators at the national level can mask severe biophysical and socio-economic trade-offs at local and regional scales (Wada et al. 2016).

There is *medium evidence* and *high agreement* that SDGs must not be pursued independently, but in a manner that recognises trade-offs and synergies with each other, consistent with a goal of ‘policy coherence.’ Policy coherence also refers to spatial trade-offs and geo-political implications within and between regions and countries implementing SDGs. For instance, supply side food security initiatives of land-based agriculture are impacting marine fisheries globally through creation of dead-zones due to agricultural run-off (Diaz and Rosenberg 2008).

SDG 7 (Affordable and clean energy) and efficient and less carbon intensive transportation (SDG 7 and 9) are important SDGs related to mitigation with adaptation co-benefits, but have local trade-offs with biodiversity and competing uses of land and rivers (see Case Study: Green Energy: Biodiversity Conservation vs Global Environment Targets) (*medium evidence, high agreement*) (Bogardi et al. 2012) (Nilsson and Berggren 2000; Hoeninghaus et al. 2009) (Winemiller et al. 2016). This has occurred despite emerging knowledge about the role that rivers and riverine ecosystems play in human development and in generating global, regional and local ecosystem services (Nilsson and Berggren 2000; Hoeninghaus et al. 2009). The transformation of river ecosystems for irrigation, hydropower and water requirements of societies worldwide is the biggest threat to fresh-water and estuarine biodiversity and

1 ecosystems services (Nilsson and Berggren 2000; Vörösmarty et al. 2010). These projects
2 address important energy and water-related demands, but their economic benefits are often
3 overestimated in relation to trade-offs with respect to food (river capture fisheries),
4 biodiversity and downstream ecosystem services (Winemiller et al. 2016). Some trade-offs
5 and synergies related to SDG7 impact aspirations of greater welfare and well-being, as
6 well as physical and social infrastructure for sustainable development (Fuso Nerini et al.
7 2018)(see 7.5.6.1 where tradeoffs exist between climate mitigation and food).

8 There are also spatial trade-offs related to large river diversion projects and export of
9 “virtual water” through water intensive crops produced in one region exported to another,
10 with implications for food-security, water security and downstream ecosystem services of
11 the exporting region (Hanasaki et al. 2010; Verma et al. 2009). Synergies include cropping
12 adaptation that increase food system production and eliminate hunger (SDG2) (Rockström et
13 al. 2017; Lipper et al. 2014a; Neufeldt et al. 2013). Well-adapted agricultural systems have
14 shown to have synergies - positive returns on investment and contribution to safe drinking
15 water, health, biodiversity and equity goals (DeClerck 2016). Assessing the water footprint
16 of different sectors at the river basin scale can provide insights for interventions and decision
17 making(Zeng et al. 2012)

18 Sometimes the trade-offs in SDGs can arise in the articulation and nested hierarchy of
19 seventeen goals and targets under them. In terms of aquatic life and ecosystems, there is an
20 explicit SDG for sustainable management of marine life (SDG 14, Life below Water). There
21 is no equivalent goal exclusively for fresh-water ecosystems, but hidden under SDG 6
22 (Clean Water and Sanitation) out of 6 listed targets, the sixth target is about protecting and
23 restoring water-related ecosystems, which suggests a lower order of global priority
24 compared to being listed as a goal in itself (e.g., SDG 14).

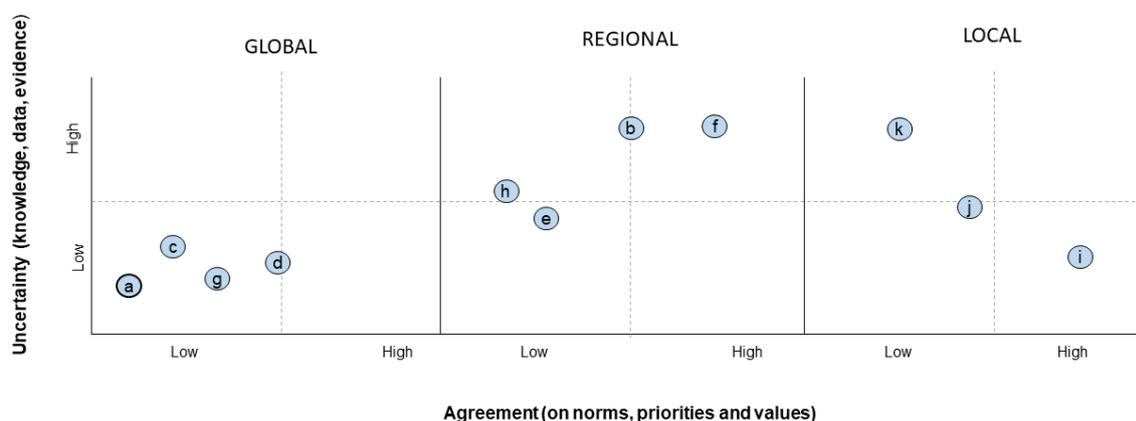
25 There is *limited evidence and limited agreement* that binary evaluations of individual SDGs
26 and synergies and trade-offs that categorise interactions as either ‘beneficial’ or ‘adverse’
27 may be subjective and challenged further by the fact that feedbacks can often not be
28 assigned as unambiguously positive or negative (Blanc et al. 2017). The Special Report on
29 Global Warming of 1.5°C notes, “A reductive focus on specific SDGs in isolation may
30 undermine the long-term achievement of sustainable climate change mitigation” (Holden et
31 al. 2017). Greater work is needed to tease out these relationships; studies that include
32 quantitative modelling (see Karnib 2017) and nuanced scoring scales (ICSU 2017) of these
33 relationships have started.

34 A nexus approach is increasingly being adopted to explore synergies and trade-offs between
35 a select subset of goals and targets (such as the interaction between water, energy, and food
36 (see, e.g., Yumkella and Yillia 2015; Conway et al. 2015; Ringler et al. 2015)). However,
37 even this approach ignores systemic properties and interactions across the system as a whole
38 (Weitz et al. 2017a). Pursuit of certain targets in one area can generate rippling effects across
39 the system, and these effects in turn can have secondary impacts on yet other targets. (Weitz
40 et al. 2017a) found that SDG target 13.2 (climate change policy/ planning) is influenced by
41 actions in six other targets. SDG 13.1 (climate change adaption) and also 2.4 (food
42 production) receive the most positive influence from progression in other targets.

43 There is *medium evidence and high agreement* that to be effective, truly sustainable, and to
44 reduce or mitigate emerging risks, SDGs need knowledge dissemination and policy
45 initiatives that recognise and assimilate concepts of co-production of ecosystem services in
46 socio-ecological systems, cross-scale linkages, uncertainty, spatial and temporal trade-offs

1 between SDGs and ecosystem services that recognise biophysical, social and political
 2 constraints and an understanding of how social change occurs at various scales (Rodríguez et
 3 al. 2006; Norström et al. 2014; Palomo et al. 2016). Several methods and tools are proposed
 4 in literature to address and understand SDG interactions. Nilsson et al. (2016a) suggest
 5 going beyond a simplistic synergies-trade-offs framing to understanding various relationship
 6 dimensions proposing a seven-point scale to understand these interactions.

7 This approach, and the identification of clusters of synergy, can help indicate that
 8 government ministries work together or establish collaborations to reach their specific goals.
 9 Finally, context specific analysis is needed. Synergies and trade-offs will depend on the
 10 natural resource base (such as land or water availability), governance arrangements,
 11 available technologies, and political ideas in a given location (Nilsson et al. 2016b). Figure
 12 7.7 below shows that at the global scale there is less uncertainty in the evidence surrounding
 13 SDGs, but also less agreement on norms, priorities and values for SDG implementation.
 14 Although there is some agreement on the regional and local scale surrounding SDGs, there is
 15 higher certainty on the science surrounding ESs.



18 **Figure 7.7 and Table 7.6: Risks at various scales, levels of uncertainty and agreement in relation to trade-**
 19 **offs among SDGs and other goals**

| | Land-climate-society Hazard | SDGs impacted or involved in mutual trade-offs | Selected Literature |
|---|--|---|--|
| a | Decline of fresh-water and riverine ecosystems | 2,3,6,7,8,12,16,18 | (Falkenmark 2001; Zarfl et al. 2014; Canonico et al. 2005) |
| b | Forest browning | 3, 8,13,15, | (Verbyla 2011; Krishnaswamy et al. 2014; McDowell and Allen 2015b; Anderegg et al. 2013; Samanta et al. 2010) |
| c | Exhaustion of ground water | 1,3,6,8,11,12,13,18 | (Barnett and O'Neill 2010; Wada et al. 2010; Harootunian 2018; Dalin et al. 2017; Rockström, Johan Steffen et al. 2009; Falkenmark 2001) |
| d | Loss of biodiversity | 6,7,12,15,18 | (Pereira et al. 2010; Pascual et al. 2017; Pecl et al. 2017; Jumani et al. 2017, 2018) |
| e | Extreme events in cities and towns | 3,6,11,13 | (Douglas et al. 2008; Stone et al. 2010; Chang et al. 2007; Hanson et al. 2011); |

| | | | |
|---|--|---------------------------|---|
| f | Stranded assets | 8, 9,11,12,13 | (Ansar et al. 2013; Chasek et al. 2015; Melvin et al. 2017; Surminski 2013; Hallegatte et al. 2013; Larsen et al. 2008; Nicholls and Cazenave 2010) |
| g | Expansion of the agricultural frontier into tropical forests | 15, 13 | (Celentano et al. 2017; Nepstad et al. 2008; Bogaerts et al. 2017; Fearnside 2015; Beuchle et al. 2015; Grecchi et al. 2014) |
| h | Food and nutrition security | 2,1,3,10, 11 | (Hasegawa et al. 2018a; Frank et al. 2017; Fujimori et al. 2018b; Zhao et al. 2017) |
| i | Emergence of Infectious Diseases | 3,1,6, 10, 11, 12, 13 | (Wu et al. 2016; Patz et al. 2004; McMichael et al. 2006; Young et al. 2017b; Smith et al. 2014a; Tjaden et al. 2017; Naicker 2011) |
| j | Decrease in Agricultural Productivity | 2,1,3,10, 11, 13 | (Porter et al. 2014; Müller et al. 2013; Rosenzweig et al. 2014) |
| k | Expansion of farm and fish ponds | 1, 2, 3, 6, 8, 10, 13, 14 | (Kale 2017; Boonstra and Hanh 2015) |

- 1
- 2 Sustainable Development Goals
- 3 1: No Poverty
- 4 2: Zero Hunger
- 5 3: Good Health and Well-being
- 6 4: Quality Education
- 7 5: Gender Equality
- 8 6: Clean Water and Sanitation
- 9 7: Affordable and Clean Energy
- 10 8: Decent Work and Economic Growth
- 11 9: Industry, Innovation and Infrastructure
- 12 10: Reduced Inequality
- 13 11: Sustainable Cities and Communities
- 14 12: Responsible Consumption and Production
- 15 13: Climate Action
- 16 14: Life Below Water
- 17 15: Life on Land
- 18 16: Peace and Justice Strong Institutions
- 19 17: Partnerships to achieve the goals

20

21

7.5.6.3. Forests and agriculture

22

23

24

25

26

27

28

29

30

31

32

33

34

35

36

37

Retaining existing forests, restoring degraded forest and afforestation are response options for climate change mitigation with adaptation benefits (6.4.1). Policies at various levels of governance that foster ownership, autonomy, and provide incentives for forest cover can reduce trade-off between carbon sinks in forests and local livelihoods (especially when the size of forest commons is sufficiently large) (Chhatre and Agrawal 2009; Locatelli et al. 2014) (see Table 7.6 this section, Case Study: Forest conservation instruments: REDD+ in the Amazon and India, 7.4.6).

Forest restoration for mitigation through carbon sequestration and other ecosystem services or co-benefits (e.g., hydrologic, NTFP, timber and tourism) can be passive or active (although both types largely exclude livestock). Passive restoration is more economically viable in relation to restoration costs as well as co-benefits in other ESs, calculated on a NPV basis, especially under flexible carbon credits (Cantarello et al. 2010). Restoration can be more cost effective with positive socioeconomic and biodiversity conservation outcomes, if costly and simplistic planting schemes are avoided (Menz et al. 2013). Passive restoration takes longer to demonstrate co-benefits and net economic gains, can be confused with land abandonment in some regions and countries, and therefore secure land-tenure at individual or community scales is important for its success (Zahawi et al. 2014). Potential approaches include improved markets and payment schemes for ecosystem services (Tengberg et al. 2016)(see

1 7.4.6). Proper targeting of incentive schemes and reducing poverty through access to ecosystem
2 services requires knowledge regarding the distribution of beneficiaries and about those whose
3 livelihoods are likely to be impacted in what manner (Nayak et al. 2014; Loaiza et al. 2015; Vira et al.
4 2012). Institutional arrangements to govern ecosystems are believed to synergistically influence
5 maintenance of carbon storage and forest based livelihoods, especially when they incorporate local
6 knowledge and decentralised decision making (Chhatre and Agrawal 2009). Earning carbon credits
7 from reforestation with native trees involves a higher cost of the certification and validation processes,
8 increasing the temptation to choose fast-growing (perhaps non-native) species with consequences for
9 native biodiversity. Strategies and policies that aggregate landowners or forest dwellers are needed to
10 reduce the cost to individuals and payment for ecosystem services (PES) schemes can generate
11 synergies (Bommarco et al. 2013; Chhatre and Agrawal 2009). Bundling several PES schemes that
12 address more than one ES can increase income generated by forest restoration (Brancalion et al.
13 2012). In the forestry sector, there is evidence that adaptation and mitigation can be fostered in
14 concert. A recent assessment of the California forest offset program shows that such programs, by
15 compensating individuals and industries for forest conservation, can deliver mitigation and
16 sustainability co-benefits (Anderson et al. 2017). Adaptive forest management focussing on re-
17 introducing native tree species can provide both mitigation and adaptation benefit by reducing fire
18 risk and increasing carbon storage (Astrup et al. 2018).

19 In the agricultural sector, there has been little published empirical work on interactions between
20 adaptation and mitigation policies. Smith and Oleson (2010) describe potential relationships,
21 focussing particularly on the arable sector and predominantly on mitigation efforts and more on
22 measures than policies. The considerable potential of the agro-forestry sector for synergies and
23 contributing to increasing resilience of tropical farming systems is discussed in (Verchot et al. 2007)
24 with examples from Africa.

25 ‘Climate Smart Agriculture’ has emerged in recent years as an approach to integrate food security and
26 climate challenges. The three pillars of CSA are to: (1) adapt and build resilience to climate change;
27 (2) reduce GHG emissions, and; (3) sustainably increase agricultural productivity, ultimately
28 delivering ‘triple-wins’ (Lipper et al. 2014c). While the concept is conceptually appealing, a range of
29 criticisms, contradictions and challenges exist in using CSA as the route to resilience in global
30 agriculture, notably around the political economy (Newell and Taylor 2017), the vagueness of the
31 definition, and consequent assimilation by the mainstream agricultural sector, as well as issues around
32 monitoring, reporting and evaluation (Arakelyan et al. 2017).

33 Land-based mitigation is facing important trade-offs with food production, biodiversity and local bio
34 geophysical effects (Humpeöder et al. 2017; Krause et al. 2017; Robledo-Abad et al. 2017; Boysen
35 et al. 2016, 2017a,b). Synergies between bio energy and food security could be achieved by investing
36 in a combination of instruments including technology and innovations, infrastructure, pricing, flex
37 crops, and improved communication and stakeholder engagement (Kline et al. 2017). Managing these
38 trade-offs might also require demand side interventions including dietary change incentives (see
39 5.7.1).

40 Synergies and trade-offs also result from interaction between policies (policy interplay (Urwin and
41 Jordan 2008)) at different levels of policy (vertical) and across different policies (horizontal) – see
42 also section on policy coherence. If policy mixes are designed appropriately, acknowledging and
43 incorporating trade-offs and synergies, they are more apt to deliver an outcome such as transitioning
44 to sustainability (Howlett and Rayner 2013; Huttunen et al. 2014) (*medium evidence and medium
45 agreement*). However, there is *medium evidence and medium agreement* that evaluating policies for
46 coherence in responding to climate change and its impacts is not occurring, and policies are instead
47 reviewed in a fragmented manner (Hurlbert and Gupta 2016).

1 In the forestry sector, there is evidence that adaptation and mitigation can be fostered in concert. A
2 recent assessment of the California forest offset program shows that such programs, by compensating
3 individuals and industries for forest conservation, can deliver mitigation and sustainability co-benefits
4 (Anderson et al. 2017). Adaptive forest management focussing on re-introducing native tree species
5 can provide both mitigation and adaptation benefit by reducing fire risk and increasing carbon storage
6 (Astrup et al. 2018).

7 Land-based mitigation is facing important trade-offs with food production, biodiversity and local bio
8 geophysical effects (Humpenöder et al. 2017; Krause et al. 2017; Robledo-Abad et al. 2017; Boysen
9 et al. 2016, 2017a,b). Synergies between bio energy and food security could be achieved by investing
10 in a combination of instruments including technology and innovations, infrastructure, pricing, flex
11 crops, and improved communication and stakeholder engagement (Kline et al. 2017). Managing these
12 trade-offs might also require demand side interventions including dietary change incentives.

14 **7.5.6.4. Water, food and aquatic ES**

15 Trade-offs between some types of water use (eg irrigation for food security) and other ecosystem
16 services are expected to intensify under climate change (Hanjra and Ejaz Qureshi 2010). There is an
17 urgency to develop approaches to understand and communicate this to policy and decision makers
18 (Zheng et al. 2016). Reducing water use in agriculture (Mekonnen and Hoekstra 2016) through
19 policies on both supply and demand side such as shift to less-water intensive crops (Richter et al.
20 2017; Fishman et al. 2015), and shift in diets (Springmann et al. 2016) has potential to reduce trade-
21 offs between food security and fresh-water aquatic ecosystem services (*medium evidence, high*
22 *agreement*). There is strong evidence that improved efficiency in irrigation can actually increase
23 overall water use in agriculture and therefore its contribution to improved flows in rivers is
24 questionable (Ward and Pulido-Velazquez 2008).

25 There are now powerful new analytical approaches, high-resolution data and decision making tools
26 that help to predict cumulative impacts of dams, assess trade-offs between engineering and
27 environmental goals, and can help funders and decision makers compare alternative sites or designs
28 for dam building as well as manage flows in regulated rivers based on experimental releases and
29 adaptive learning. This could minimise ecological costs and maximise synergies with other
30 development goals under climate change (Poff et al. 2003; Winemiller et al. 2016). Furthermore the
31 adoption of metrics based on the emerging concept of Nature’s Contributions to People (NCP) under
32 the IPBES framework brings in non-economic instruments and values that in combination with
33 conventional valuation of ecosystem services approaches could elicit greater support for non-
34 consumptive water use of rivers for achieving SDG goals (De Groot et al. 2010; Pascual et al. 2017).

36 **7.5.6.5. Considering Synergies and Tradeoffs to Avoid Maladaptation**

37 Coherent policies that consider synergies and tradeoffs can also reduce the likelihood of
38 maladaptation, which is the opposite of sustainable adaptation (Magnan et al. 2016). Sustainable
39 adaptation is adaptation that “contributes to socially and environmentally sustainable development
40 pathways including both social justice and environmental integrity” (Eriksen et al. 2011). In AR5
41 there was *medium evidence* and *high agreement* that maladaptation is ‘a cause of increasing concern
42 to adaptation planners, where intervention in one location or sector could increase the vulnerability of
43 another location or sector, or increase the vulnerability of a group to future climate change’ (Noble et
44 al. 2014). AR5 recognised that maladaptation arises not only from inadvertent, badly planned
45 adaptation actions, but also from deliberate decisions where wider considerations place greater
46 emphasis on short-term outcomes ahead of longer-term threats, or that discount, or fail to consider,
47 the full range of interactions arising from planned actions (Noble et al. 2014).

1 Some maladaptations are only beginning to be recognised as we become aware of unintended
2 consequences of decisions. An example prevalent across many countries is irrigation as an adaptation
3 to water scarcity. During a drought from 2007–2009 in California, farmers adapted by using more
4 groundwater thereby depleting groundwater elevation by 15 metres. This volume of groundwater
5 depletion is unsustainable environmentally and also emits GHG emissions during the pumping
6 (Christian-Smith et al. 2015). Despite the three years of drought, the agricultural sector performed
7 financially well, due to the groundwater use and crop insurance payments. Drought compensation
8 programmes through crop insurance policies may reduce the incentive to shift to lower water-use
9 crops, thereby perpetuating the maladaptive situation. Another example of maladaptation that may
10 appear as adaptation to drought is pumping out groundwater and storing in surface farm ponds with
11 consequences for water justice, inequity and sustainability (Kale 2017). These examples highlights
12 both the potential for maladaptation from farmers’ adaptation decisions as well as the unintended
13 consequences of policy choices and illustrates the findings of Barnett and O’Neill (2010) that
14 maladaptation can include high opportunity costs (including economic, environmental, and social);
15 reduced incentives to adapt (adaptation measures that reduce incentives to adapt by not addressing
16 underlying causes); and path dependency or trajectories that are difficult to change.

17 In practice, maladaptation is a specific instance of policy incoherence, and it may be useful to develop
18 a framework in designing policy to avoid this type of trade-off. This would specify the type, aim and
19 target audience of an adaptation action, decision, project, plan, or policy designed initially for
20 adaptation, but actually at high risk of inducing adverse effects either on the system in which it was
21 developed, or another connected system, or both. The assessment requires identifying system
22 boundaries including temporal and geographical scales at which the outcome are assessed (Magnan
23 2014; Juhola et al. 2016). National level institutions that cover the spectrum of sectors affected, or
24 enhanced collaboration between relevant institutions is expected to increase the effectiveness of
25 policy instruments, as are joint programmes and funds (Morita and Matsumoto 2018).

26 As new knowledge about trade-offs and synergies amongst land-climate processes emerges regionally
27 and globally, concerns over emerging risks and the need for planning policy responses grow. There is
28 *medium evidence and medium agreement* that trade-offs currently do not figure into existing climate
29 policies including NDCs and SDGs being vigorously pursued by some countries (Woolf et al. 2018).
30 For instance, the biogeophysical co-benefits of reduced deforestation and re/afforestation measures
31 (Chapter 6) are usually not accounted for in current climate policies or in the NDCs, but there is
32 increasing scientific evidence to include them as part of the policy design (Findell et al. 2017; Hirsch
33 et al. 2018; Bright et al. 2017).

34

35 **Case Study: Green Energy: Biodiversity Conservation vs Global Environment Targets?**

36

37 Green and renewable energy and transportation are emerging as an important part of climate change
38 mitigation globally (*medium evidence, high agreement*) (McKinnon 2010; Zarfl et al. 2015; Creutzig
39 et al. 2017). Evidence is however emerging across many biomes (from coastal to semi-arid and
40 humid) how green energy may have significant trade-offs with biodiversity and ecosystem services
41 thus demonstrating the need for closer environmental scrutiny and safeguards (Gibson et al.
42 2017)(Hernandez et al. 2015). In most cases, the accumulated impact of pressures from decades of
43 land-use and habitat loss set the context within which the potential impacts of renewable energy
44 generation need to be considered.

45

46 Small hydropower or SHPs were until recently considered as environmentally benign compared to
47 large dams and are poorly understood, especially since the impacts of clusters of small dams are just
48 becoming evident (Mantel et al. 2010; Fencel et al. 2015; Kibler and Tullos 2013). SHPs (<25/30 MW)
49 and being labelled “green” are often exempt from environmental scrutiny (Abbasi and Abbasi 2011;

1 Pinho et al. 2007; Premalatha et al. 2014b; Era Consultancy 2006). Being promoted in mountainous
2 global biodiversity hotspots, SHPs have changed the hydrology, water quality and ecology of head-
3 water streams and neighbouring forests significantly. SHPs have created dewatered stretches of
4 stream immediately downstream and introduced sub-daily to sub-weekly hydro-pulses that have
5 transformed the natural dry-season flow regime. Hydrologic and ecological connectivity have been
6 impacted, especially for endemic fish communities and fragmented forests in the Himalayas and
7 Western Ghats biodiversity hotspots in India, and regions in China, and Central America (*medium*
8 *evidence, medium agreement*) (Jumani et al. 2017, 2018; Chhatre and Lakhanpal 2018; Anderson et
9 al. 2006; Grumbine and Pandit 2013). Some regions have opposed SHPs over concerns about impacts
10 on local culture and livelihoods (Jumani et al. 2017, 2018; Chhatre and Lakhanpal 2018).

11 Large scale solar farms that involve large land resources are being installed at a rapid rate. In India,
12 semi-arid and arid regions are targeted for wind and solar farms. India's renewable energy targets
13 are often sited in semi-arid areas which includes the last remaining habitats of the highly
14 endangered Great Indian Bustard (*Ardeotis nigriceps*). Installing solar and wind farms linked to
15 lethal power transmission lines cause mortality of a species whose global population is now reduced
16 to about 150 (Collar et al. 2015). The loss of habitat over the decades has been largely due to
17 agricultural intensification driven by irrigation and bad management in designated reserves (Collar
18 et al. 2015; Ledec, George C.; Rapp, Kennan W.; Aiello 2011) but intrusion of power lines in its
19 last remaining refuges is a major worry for its future persistence (Government of India 2012). In
20 many regions around the world, wind-turbines and solar farms pose a threat to many other species
21 especially predatory birds and insectivorous bats (*medium evidence, medium agreement*) (Thaker,
22 M, Zambre, A. Bhosale 2018) and disrupt habitat connectivity (Northrup and Wittemyer 2013).

23 Additionally, conversion of rivers into waterways has been touted as a fuel-efficient (low carbon
24 emitting) and environment-friendly alternative to surface land transport (IWAI 2016; Dharmadhikary,
25 S., and Sandbhor 2017). India's National Waterways (funded partly by a USD 375 million loan from
26 the World Bank) seeks to cut transportation time and costs and reduce carbon emissions from road
27 transport (Admin 2017). However given the low water levels in India's rivers in the dry-season
28 (due to upstream demands and abstraction) the programme relies on large scale dredging to
29 maintain deep channels. Evidence from elsewhere suggests that dredging could severely impact the
30 water quality, human health and habitat of fish species (Junior et al. 2012; Martins et al. 2012),
31 disrupt artisanal fisheries and potentially cause severe threat to the endangered Ganges River
32 Dolphin (*Platanista gangetica*), India's National Aquatic Animal (Kelkar 2016). The most severe
33 impact of dredging and vessel traffic on this unique species is the disruption through under-water
34 noise of the acoustic signals that the endangered and naturally blind animal relies on for navigation,
35 foraging and communication (*low evidence, medium agreement*) (Dey Mayukh 2018). Off-shore
36 renewable energy projects in coastal zones have been known to have similar impacts on marine fauna
37 (Gill 2005).

38 Policy response to mitigate and reduce the negative impacts of small dams include changes in SHP
39 operations and policies to enable the conservation of river fish diversity. These include mandatory
40 environmental impact assessments, conserving remaining undammed headwater streams in regulated
41 basins, maintaining adequate environmental flows, and implementing other adaptation measures
42 based on experiments with active management of fish communities in impacted zones (Jumani et al.
43 2018). Location of large solar farms needs to be carefully scrutinised (Sindhu et al. 2017). For
44 mitigating negative impacts of power lines associated with solar and wind-farms in bustard habitat,
45 suggested measures include diversion structures to prevent collision, underground cables and
46 avoidance in core wildlife habitat as well as incentives for maintaining low intensity rain-fed
47 agriculture and pasture around existing reserves, and curtailing harmful infrastructure in priority areas
48 (Collar et al. 2015). Mitigation for minimising the ecological impact of Inland Waterways on
49 biodiversity and fisheries is more complicated but may involve improved boat technology to reduce
50 under-water noise, maintaining ecological flows and thus reduced dredging, and avoidance in key
51 habitats (Dey Mayukh 2018).

1 The management of ecological trade-offs of green energy and green infrastructure and transportation
2 projects may be crucial for long-term sustainability and acceptance of emerging low-carbon
3 economies.

5 **7.6. Governance: Governing the land-climate interface**

6 Building on the definition of governance in section 7.1.2, governance situates decision making and
7 selection or calibration of policy instruments within the reality of the multitude of actors operating in
8 respect of land and climate interactions. Governance includes all of the processes, structures, rules and
9 traditions that govern; governance processes may be undertaken by actors including a government,
10 market, organisation, or family (Bevir 2011). Governance processes determine how people in
11 societies make decisions (Patterson et al. 2017) and involve the interactions among formal and
12 informal institutions (see 7.4.1) through which people articulate their interests, exercise their legal
13 rights, meet their legal obligations, and mediate their differences (Plummer and Baird 2013).

14 The act of governance “is a social function centred on steering collective behaviour toward desired
15 outcomes and away from undesirable outcomes” (Young 2017a), here sustainable climate resilient
16 development. This definition of governance allows for it to be decoupled from the more familiar
17 concept of government and studied in the context of complex human-environment relations and
18 environmental and resource regimes (Young 2017a) and used to address the interconnected
19 challenges facing food and agriculture (FAO 2017b). These challenges include assessing, combining,
20 and implementing policy instruments at different governance levels in a mutually reinforcing way,
21 managing trade-offs while capitalising on synergies (see 7.5.6), and employing experimentalist
22 approaches for improved and effective governance (FAO 2017b), here adaptive climate governance
23 (7.6.3). Emphasising governance also represents a shift of traditional resource management (focused
24 on hierarchical state control) towards recognition that political and decision making authority can be
25 exercised through interlinked groups of diverse actors (Kuzdas et al. 2015).

26 This section will start with describing institutions and institutional arrangements (the core of a
27 governance system (Young 2017)) that build adaptive and mitigative capacity, outlining modes, levels
28 and scales of governance for sustainable climate resilient development, describing adaptive climate
29 governance that responds to uncertainty, exploring institutional dimensions of adaptive governance
30 that create an enabling environment for strong institutional capital, discussing land tenure (an
31 important institutional context for effective and appropriate selection of policy instruments), and end
32 with the participation of people in decision making through inclusive governance.

34 **7.6.1. Institutions Building Adaptive and Mitigative Capacity**

35 Institutions are rules and norms held in common by social actors that guide, constrain, and shape
36 human interaction. Institutions can be formal, such as laws, policies, and structured decision making
37 processes (see 7.5.1.1) or informal, such as norms, conventions, and decision making following
38 customary norms and habits (see 7.5.1.2). Organisations – such as parliaments, regulatory agencies,
39 private firms, and community bodies – as well as people, develop and act in response to institutional
40 frameworks and the incentives they frame. “Institutions can guide, constrain, and shape human
41 interaction through direct control, through incentives, and through processes of socialization” (AR5,
42 2014 at p. 1768). Nations with “well developed institutional systems are considered to have greater
43 adaptive capacity,” and better institutional capacity to help deal with risks associated with future
44 climate change (IPCC, 2001 at p. 896). Institutions may also prevent the development of adaptive

1 capacity when they are ‘sticky’ or characterised by strong path dependence (Mahoney 2000) (North
2 1991) and prevent changes that are important to address climate change (see 7.4.9).

3 Formal and informal governance structures are composed of these institutionalised rule systems that
4 determine vulnerability as they influence power relations, risk perceptions and establish the context
5 wherein risk reduction, adaptation and vulnerability are managed (Cardona 2012). Governance
6 institutions determine the management of a community’s assets, the community members’
7 interrelationship, and their relationships with natural resources (Hurlbert and Diaz 2013). Traditional
8 or locally-evolved institutions, backed by cultural norms, can contribute to resilience and adaptive
9 capacity. Anderson et al. suggest these are particularly a feature of dry land societies that are highly
10 prone to environmental risk and uncertainty (Anderson et al. 2010). Concepts of resilience, and
11 specifically the resilience of socio-ecological systems have advanced analysis of adaptive institutions
12 and adaptive governance in relation to climate change and land (Boyd and Folke 2011a). In their
13 characterisation, “resilience is the ability to reorganise following crisis, continuing to learn, evolving
14 with the same identity and function, and also innovating and sowing the seeds for transformation. It is
15 a central concept of adaptive governance” (Boyd and Folke 2012). In the context of complex and
16 multi-scale socio-ecological systems, important features of adaptive institutions that contribute to
17 resilience include the characteristics of an adaptive governance system (see 7.6.6).

18 There is *high confidence* that adaptive institutions include a strong learning dimension and include:

- 19 (1) Institutions advancing the capacity to learn through availability, access to, accumulation of,
20 and interpretation of information (such as drought projections, costing of alternatives land,
21 food, and water strategies). Government supported networks, learning platforms, and
22 facilitated interchange between actors with boundary and bridging organisations, creates the
23 necessary self-organisation to prepare for the unknown. Through transparent, flexible
24 networks, whole sets of complex problems of land, food, and climate can be tackled to
25 develop shared visions and critique land and food management systems assessing gaps and
26 generating solutions;
- 27 (2) Institutions advancing learning by experimentation (in interpretation of information, new
28 ways of governing, and treating policy as an ongoing experiment) through many interrelated
29 decisions, but especially those that connect the social to the ecological and entail anticipatory
30 planning by considering a longer term time frame. Mechanisms to do so include ecological
31 stewardship and rituals and beliefs of indigenous societies that sustain ecosystem services;
- 32 (3) Institutions that decide on pathways to realise system change through cultural, inter and intra
33 organisational collaboration, with a flexible regulatory framework allowing for new cognitive
34 frames of ‘sustainable’ land management and ‘safe’ water supply that open alternative
35 pathways (Karpouzoglou et al. 2016; Bettini et al. 2015; Boyd et al. 2015; Boyd and Folke
36 2011b)) (Boyd and Folke 2012).

37 Shortcomings of resilience theory include limits in relation to its conceptualisation of social change
38 (Cote and Nightingale 2012), its potential to be used as a normative concept implying politically
39 prescriptive policy solutions (Thorén and Olsson 2017; Weichselgartner and Kelman 2015; Milkoreit
40 et al. 2015), its applicability to local needs and experiences (Forsyth 2018), and its potential to hinder
41 evaluation of policy effectiveness (Newton 2016; Olsson et al. 2015b). Regardless, concepts of
42 adaptive institutions building adaptive capacity in complex socio-ecological systems governance
43 have progressed (Karpouzoglou et al. 2016; Dwyer and Hodge 2016) in relation to adaptive
44 governance (Koontz et al. 2015).

45 The study of institutions of governance, levels, modes, and scale of governance, in a multi-level and
46 polycentric fashion is important because of the multi-scale nature of the challenges to resilience,
47 dissemination of ideas, networking and learning.

7.6.2. Integration - Levels, Modes, and Scale of Governance for Sustainable Development

Different types of governance can be distinguished according to intended levels (e.g., local, regional, global), domains (national, international, transnational), modes (market, network, hierarchy), and scales (global regimes to local community groups) (Jordan et al. 2015b). Implementation of climate change adaptation and mitigation has been impeded by institutional barriers including multi-level governance and policy integration issues (Biesbroek et al. 2010). To overcome these barriers, climate governance has evolved significantly beyond the national and multilateral domains that tended to dominate climate efforts and initiatives during the early years of the UNFCCC. The climate challenge has been placed in an “earth system” context, showing the existence of complex interactions and governance requirements across different levels and calling for a radical transformation in governance, rather than minor adjustments (Biermann et al. 2012). Climate governance literature has expanded since AR5 in relation to the sub-national and transnational levels, but all levels and their interconnection is important. Expert thinking has evolved from implementing good governance at high levels of governance (with governments) to a decentred problem solving approach consistent with adaptive governance. This approach involves iterative bottom up and experimental mechanisms that might entail addressing tenure of land or forest management through a territorial approach to development, thereby supporting multi-sectoral governance in local, municipal, and regional contexts (FAO 2017b).

Local action in relation to mitigation and adaptation continues to be important by complementing and advancing global climate policy (Ostrom 2012). Sub-national governance efforts for climate policy, especially at the level of cities and communities, have become significant during the past decades (*medium evidence, medium agreement*) (Castán Broto 2017; Floater et al. 2014; Albers et al. 2015; Archer et al. 2014). A transformation of sorts has been underway through deepening engagement from the private sector and NGOs as well as Government involvement at multiple levels. It is now recognised that business organisations, civil society groups, citizens, and formal governance all have important roles in governance for sustainable development (Kemp et al. 2005).

Transnational governance efforts have increased in number, with application across different economic sectors, geographical regions, civil society groups and non-governmental organisations. When it comes to climate mitigation, transnational mechanisms generally focus on networking and may not necessarily be effective in terms of promoting real emissions reductions (Michaelowa and Michaelowa 2017). However, acceleration in national mitigation measures has been determined to coincide with landmark international events such as the build up to the Copenhagen Climate Conference (Iacobuta et al. 2018). There is a tendency for transnational governance mechanisms to lack monitoring and evaluation procedures (Jordan et al. 2015a).

To address shortcomings of transnational governance, polycentric governance considers the interaction between actors at different levels of governance (local, regional, national, and global) for a more nuanced understanding of the variation in diverse governance outcomes in the management of common-pool resources (such as forests) based on the needs and interests of citizens (Nagendra and Ostrom 2012). A more “polycentric climate governance” system has emerged that incorporates bottom-up initiatives that can support and synergise with national efforts and international regimes (Ostrom 2010). Although it is clear that many more actors and networks are involved, the effectiveness of a more polycentric system remains unclear (Jordan et al. 2015a).

There is *high confidence* that a hybrid form of governance combining the advantages of centralised governance (with coordination, stability, compliance) with those of more horizontal structures (that allow flexibility, autonomy for local decision making, multi-stakeholder engagement, co-management) is required for effective mainstreaming of mitigation and adaptation in sustainable land

1 and forest management (Keenan 2015; Gupta 2014; Williamson and Nelson 2017; Liniger et al.
2 2019). Polycentric institutions self-organise developing collective solutions to local problems as they
3 arise (Koontz et al. 2015). The public sector (governments and administrative systems) are still
4 important in climate change initiatives as these actors retain the political will to implement and make
5 initiatives work (Biesbroek et al. 2018).

6 Sustainable development hinges on the holistic integration of interconnected land and climate issues,
7 sectors, levels of government, and policy instruments (see Policy Coherence 7.4.8), that address the
8 increasing volatility in oscillating systems and weather patterns (Young 2017b; Kemp et al. 2005).
9 Climate adaptation and mitigation goals must be integrated or mainstreamed into existing governance
10 mechanisms around key land use sectors such as forestry and agriculture. In the EU, mitigation has
11 generally been well-mainstreamed in regional policies but not adaptation (Hanger et al. 2015).
12 Climate change adaptation has been impeded by institutional barriers including the inherent
13 challenges of multi-level governance and policy integration (Biesbroek et al. 2010).

14 Integrative polycentric approaches to land use and climate interactions take different forms and
15 operate with different institutions and governance mechanisms. Integrative approaches can provide
16 coordination and linkages to improve effectiveness and efficiency and minimise conflicts (*high*
17 *confidence*). Different types of integration with special relevance for the land-climate interface can be
18 characterised as follows:

- 19 1. Cross-level integration: local and national level efforts must be coordinated with national and
20 regional policies and also be capable of drawing direction and financing from global regimes,
21 thus requiring multi-level governance. Integration of sustainable land management to prevent,
22 reduce, and restore degraded land is advanced with national and subnational policy includes
23 passing the necessary laws establishing frameworks and providing financial incentives.
24 Examples include: integrated territorial planning addressing specific land use decisions; local
25 landscape participatory planning with farmer associations, microenterprises, and local
26 institutions identifying hot spot areas, identifying land use pressures and scaling out
27 sustainable land management response options (Liniger et al. 2019).
- 28 2. Cross-sectoral integration: rather than approach each application or sector (e.g., energy,
29 agriculture, forestry) separately, there is a conscious effort at co-management and
30 coordination in policies and institutions, such as with the energy-water-food nexus (Biggs et
31 al. 2015).
- 32 3. End-use/market integration: often involves exploiting economies of scope across products,
33 supply chains, and infrastructure (Nuhoff-Isakhanyan et al. 2016; Ashkenazy et al. 2017). For
34 instance land-use transport models consider land use, transportation, city planning , and
35 climate mitigation (Ford et al. 2018).
- 36 4. Landscape integration: rather than physical separation of activities (e.g., agriculture, forestry,
37 grazing), uses are spatially integrated by exploiting natural variations while incorporating
38 local and regional economies (Harvey et al. 2014a). In an assessment of 166 initiatives in 16
39 countries, integrated landscape initiatives were found to address the drivers of agriculture,
40 ecosystem conservation, livelihood preservation and institutional coordination. However,
41 such initiatives struggled to move from planning to implementation due to lack of government
42 and financial support and powerful stakeholders sidelining the agenda (Zanzanaini et al.
43 2017). Special care helps ensure initiatives don't exacerbate socio-spatial inequalities across
44 diverse developmental and environmental conditions (Anguelovski et al. 2016b). Integrated
45 land use planning coordinated through multiple government levels balances property rights,
46 wildlife and forest conservation, encroachment of settlements and agricultural areas and can
47 reduce conflict (*high confidence*) (Metternicht 2018). Land use planning can also enhance
48 management of areas prone to natural disasters such as floods and resolve issues of competing
49 land uses and land tenure conflicts (Metternicht 2018).

1
2 Another way to analyse or characterise governance approaches or mechanisms might be according to
3 a temporal scale with respect to relevant events, for example those that may occur gradually vs.
4 abruptly (Cash et al. 2006). Desertification and land degradation are drawn-out processes that occur
5 over many years, whereas extreme events are abrupt and require immediate attention. Similarly, the
6 frequency of events might be of special interest, for example events that occur periodically vs. those
7 that occur infrequently and/or irregularly. In the case of food security abrupt and protracted events of
8 food insecurity might occur. There is a distinction between “hunger months” and longer-term food
9 insecurity. Some indigenous practices already incorporate hunger months whereas structural food
10 deficits have to be addressed differently (Bacon et al. 2014). Governance mechanisms that facilitate
11 rapid response to crises are quite different from those aimed at monitoring slower changes and
12 responding with longer-term measures.

13 **Governance Case Study: Biofuels and bioenergy**

14 New policies and initiatives during the past decade or so have increased support for bioenergy as a
15 non-intermittent (stored) renewable with wide geographic availability that is cost-effective in a range
16 of applications. Significant upscaling of bioenergy requires dedicated (normally land-based) sources
17 in addition to use of wastes and residues. As a result a disadvantage is high land use intensity
18 compared to other renewables (Fritsche et al. 2017b) that in turn place greater demands on
19 governance. Bioenergy, especially traditional fuels currently provides the largest share of renewable
20 energy globally and has a significant role in nearly all climate stabilisation scenarios, although
21 estimates of its potential vary widely (see Cross-Chapter Box 7 on Bioenergy and BECCS in Chapter
22 6). Policies and governance for bioenergy systems and markets must address diverse applications and
23 sectors across levels from local to global; here we briefly review the literature in relation to
24 governance for **modern** bioenergy and biofuels with respect to land and climate impacts whereas
25 **traditional biomass** use (see Glossary) (> 50% of energy used today with greater land use and GHG
26 emissions impacts in low and medium-income countries (Bailis et al. 2015; Masera et al. 2015; Bailis
27 et al. 2017a; Kiruki et al. 2017b)) is addressed elsewhere (see sections 4.5.4 and 7.4.6.4 and Cross-
28 Chapter Box 12 on Traditional Biomass in this chapter). The bioenergy cycle is relevant in accounting
29 for—and attributing—land impacts and GHG emissions (see section 2.5.1.5). Integrated responses
30 across different sectors can help to reduce negative impacts and promote sustainable development
31 opportunities (Table 6.9, Table 6.58). It is *very likely* that bioenergy expansion at a scale that
32 contributes significantly to global climate mitigation efforts (see Cross-Chapter Box 7 on Bioenergy
33 and BECCS in Chapter 6) will result in substantial land use change (Berndes et al. 2015; Popp et al.
34 2014a; Wilson et al. 2014; Behrman et al. 2015; Richards et al. 2017; Harris et al. 2015; Chen et al.
35 2017a). There is *medium evidence and high agreement* that land use change at such scale presents a
36 variety of positive and negative socio-economic and environmental impacts that lead to risks and
37 trade-offs that must be managed or governed across different levels (Pahl-Wostl et al. 2018a; Kurian
38 2017; Franz et al. 2017; Chang et al. 2016; Larcom and van Gevelt 2017; Lubis et al. 2018; Alexander
39 et al. 2015b; Rasul 2014; Bonsch et al. 2016; Karabulut et al. 2018; Mayor et al. 2015). There is
40 *medium evidence and high agreement* that impacts vary considerably with factors such as initial land
41 use type, choice of crops, initial carbon stocks, climatic region, soil types and the management regime
42 and technologies adopted (Qin et al. 2016; Del Grosso et al. 2014; Popp et al. 2017; Davis et al. 2013;
43 Mello et al. 2014; Hudiburg et al. 2015; Carvalho et al. 2016; Silva-Olaya et al. 2017; Whitaker et al.
44 2018; Alexander et al. 2015b);

45 There is *medium evidence and high agreement* that significant socio-economic impacts requiring
46 additional policy responses can occur when agricultural lands and/or food crops are used for
47 bioenergy due to competition between food and fuel (Harvey and Pilgrim 2011; Rosillo Callé and
48 Johnson 2010b), including impacts on food prices (Martin Persson 2015; Roberts and Schlenker 2013;

1 Borychowski and Czyżewski 2015; Koizumi 2014; Muratori et al. 2016; Popp et al. 2014b; Araujo
2 Enciso et al. 2016) and impacts on food security (Popp et al. 2014b; Bailey 2013; Pahl-Wostl et al.
3 2018b; Rulli et al. 2016; Yamagata et al. 2018; Kline et al. 2017; Schröder et al. 2018; Franz et al.
4 2017; Mohr et al. 2016). Additionally crops such as sugar-cane which are water-intensive when used
5 for ethanol production have a trade-off with water and downstream ecosystem services and other
6 crops more important for food security (Rulli et al. 2016; Gheewala et al. 2011). Alongside negative
7 impacts that might fall on urban consumers (who purchase both food and energy), there is *medium*
8 *evidence and medium agreement* that rural producers or farmers can increase income or strengthen
9 livelihoods by diversifying into biofuel crops that have an established market (Maltsoglou et al. 2014;
10 Mudombi et al. 2018a; Gasparatos et al. 2018a,b; von Maltitz et al. 2018; Gasparatos et al. 2018c;
11 Kline et al. 2017; Rodríguez Morales and Rodríguez López 2017; Dale et al. 2015; Lee and Lazarus
12 2013; Rodríguez-Morales 2018). A key governance mechanism that has emerged in response to such
13 concerns, especially during the past decade are standards and certification systems that include food
14 security and land rights in addition to general criteria or indicators related to sustainable use of land
15 and biomass (see section 7.4.6.3 on Standards and Certification). There is *medium evidence and*
16 *medium agreement* that policies promoting use of wastes and residues, the use of non-edible crops
17 and/or reliance on degraded and marginal lands for bioenergy could reduce land competition and
18 associated risk for food security (Manning et al. 2015; Maltsoglou et al. 2014; Zhang et al. 2018a; Gu
19 and Wylie 2017; Kline et al. 2017; Schröder et al. 2018; Suckall et al. 2015; Popp et al. 2014a; Lal
20 2013).

21 There is *medium evidence and high agreement* that good governance, including policy coherence and
22 coordination across the different sectors involved (agriculture, forestry, livestock, energy, transport)
23 (see 7.6.2) can help to reduce the risks and increase the co-benefits of bioenergy expansion
24 (Makkonen et al. 2015; Di Gregorio et al. 2017; Schut et al. 2013; Mukhtarov et al.; Torvanger 2019a;
25 Müller et al. 2015; Nkonya et al. 2015; Johnson and Silveira 2014a; Lundmark et al. 2014; Schultz et
26 al. 2015; Silveira and Johnson 2016; Giessen et al. 2016b; Stattman et al. 2018b; Bennich et al.
27 2017b). There is *medium evidence and high agreement* that the nexus approach can help to address
28 interconnected biomass resource management challenges and entrenched economic interests, as well
29 to leverage synergies in the systemic governance of risk. (Bizikova et al. 2013; Rouillard et al. 2017;
30 Pahl-Wostl 2017a; Lele et al. 2013; Rodríguez Morales and Rodríguez López 2017; Larcom and van
31 Gevelt 2017; Pahl-Wostl et al. 2018a; Rulli et al. 2016; Rasul and Sharma 2016; Weitz et al. 2017b;
32 Karlberg et al. 2015).

33 A key issue for governance of biofuels and bioenergy, as well as land use governance more generally,
34 during the past decade is the need for new governance mechanisms across different levels as land use
35 policies and bioenergy investments are scaled up and result in wider impacts (see section 7.6). There
36 is *low evidence and medium agreement* that hybrid governance mechanisms can promote sustainable
37 bioenergy investments and land use pathways. This hybrid governance can include multi-level,
38 transnational governance, and private-led or partnership-style (polycentric) governance
39 complementing national-level, strong public coordination (government and public
40 administration){7.6.2} (Pahl-Wostl 2017a; Pacheco et al. 2016; Winickoff and Mondou 2017;
41 Nagendra and Ostrom 2012; Jordan et al. 2015a; Djalante et al. 2013; Purkus, Alexandra; Gawel,
42 Erik; Thrän 2012; Purkus et al. 2018; Stattman et al.; Rietig 2018; Cavicchi et al. 2017; Stupak et al.
43 2016; Stupak and Raulund-Rasmussen 2016; Westberg and Johnson 2013; Giessen et al. 2016b;
44 Johnson and Silveira 2014b; Stattman et al. 2018b; Mukhtarov et al.; Torvanger 2019b).

45

46

Cross-Chapter Box 12: Traditional biomass use: land, climate and development implications

Francis X. Johnson (Sweden), Fahmuddin Agus (Indonesia), Rob Bailis (The United States of America.), Suruchi Bhadwal (India), Annette Cowie (Australia), Tek Sapkota (Nepal)

Introduction and significance

Most biomass used for energy today is in traditional forms (fuelwood, charcoal, agricultural residues) for cooking and heating by some 3 billion persons worldwide (IEA 2017). Traditional biomass has high land and climate impacts, with significant harvesting losses, GHG emissions, soil impacts and high conversion losses (Cutz et al. 2017b; Masera et al. 2015; Ghilardi et al. 2016a; Bailis et al. 2015; Fritsche et al. 2017b; Mudombi et al. 2018b). In addition to these impacts, indoor air pollution from household cooking is a leading cause of mortality in low and medium-income countries and affects especially women and children (Smith et al. 2014a; HEI/IHME 2018; Goldemberg et al. 2018b). In rural areas, the significant time needed for gathering fuelwood imposes further costs on women and children (Njenga and Mendum 2018; Gurung and Oh 2013a; Behera et al. 2015a).

Both agricultural and woody biomass can be upgraded and used sustainably through improved resource management and modern conversion technologies, providing much greater energy output per unit of biomass (Cutz et al. 2017b; Hoffmann et al. 2015a; Gurung and Oh 2013b). More relevant than technical efficiency is the improved quality of energy services: with increasing income levels and/or access to technologies, households transition over time from agricultural residues and fuelwood to charcoal and then to gaseous or liquid fuels and electricity (Leach 1992; Pachauri and Jiang 2008; Goldemberg and Teixeira Coelho 2004; Smeets et al. 2012a). However, most households use multiple stoves and/or fuels at the same time, known as “fuel stacking” for economic flexibility and also for sociocultural reasons (Ruiz-Mercado and Masera 2015a; Cheng and Urpelainen 2014; Takama et al. 2012).

Urban and rural use of traditional biomass

In rural areas, fuelwood is often gathered at no cost to the user and burned directly whereas in urban areas, traditional biomass use may often involve semi-processed fuels, particularly in sub-Saharan Africa where charcoal is the primary urban cooking fuel. Rapid urbanisation and/or commercialisation drives a shift from fuelwood to charcoal, which results in significantly higher wood use (*very high confidence*) due to losses in charcoal supply chains and the tendency to use whole trees for charcoal production (Santos et al. 2017; World Bank. 2009a; Hojas-Gascon et al. 2016a; Smeets et al. 2012b). One study in Myanmar found that charcoal required 23 times the land area of fuelwood (Win et al. 2018). In areas of woody biomass scarcity, animal dung and agricultural residues as well as lower quality wood are often used (Kumar Nath et al. 2013a; Go et al. 2019a; Jagger and Kittner 2017; Behera et al. 2015b). The fraction of woody biomass harvested that is not “demonstrably renewable” is the fraction of non-renewable biomass (fNRB) under UNFCCC accounting; default values for fNRB for least developed countries and small island developing states ranged from 40% to 100% (CDM Executive Board 2012). Uncertainties in woodfuel data, complexities in spatiotemporal woodfuel modelling and rapid forest regrowth in some tropical regions present sources of variation in such estimates, and some fNRB values are *likely* to have been over-estimated (McNicol et al. 2018a; Ghilardi et al. 2016b; Bailis et al. 2017b).

GHG emissions and traditional biomass

Due to overharvesting, incomplete combustion and the effects of short-lived climate pollutants, traditional woodfuels (fuelwood and charcoal) contribute 1.9-2.3% of global GHG emissions; non-renewable biomass is concentrated especially in “hotspot” regions of East Africa and South Asia

1 (Bailis et al. 2015). The estimate only includes woody biomass and does not account for possible
2 losses in soil carbon or the effects of nutrient losses from use of animal dung, which can be significant
3 in some cases (Duguma et al. 2014a; Achat et al. 2015a; Sánchez et al. 2016). Reducing emissions of
4 black carbon alongside GHG reductions offers immediate health co-benefits (Shindell et al. 2012;
5 Pandey et al. 2017; Weyant et al. 2019a; Sparrevik et al. 2015). Significant GHG emissions
6 reductions, depending on baseline or reference use, can be obtained through fuel-switching to gaseous
7 and liquid fuels, sustainable harvesting of woodfuels, upgrading to efficient stoves, and adopting
8 high-quality processed fuels such as wood pellets (*medium evidence, high agreement*) (Wathore et al.
9 2017; Jagger and Das 2018; Quinn et al. 2018a; Cutz et al. 2017b; Carter et al. 2018; Bailis et al.
10 2015; Ghilardi et al. 2018; Weyant et al. 2019b; Hoffmann et al. 2015b).

11 **Land and forest degradation**

12 Land degradation is itself a significant source of GHG emissions and biodiversity loss, with
13 overharvesting of woodfuel as a major cause in some regions and especially in sub-Saharan Africa
14 (Pearson et al. 2017; Joana Specht et al. 2015a; Kiruki et al. 2017b; Ndegwa et al. 2016; McNicol et
15 al. 2018b). Reliance on traditional biomass is quite land-intensive: supplying one household
16 sustainably for a year can require more than half a hectare of land, which, in dryland countries such as
17 Kenya, can result in substantial percentage of total tree cover (Fuso Nerini et al. 2017). In sub-
18 Saharan Africa and in some other regions, land degradation is widely associated with charcoal
19 production (*high confidence*), often in combination with timber harvesting or clearing land for
20 agriculture (Kiruki et al. 2017a; Ndegwa et al. 2016; Hojas-Gascon et al. 2016b). Yet charcoal makes
21 a significant contribution to livelihoods in many areas and thus in spite of the ecological damage,
22 halting charcoal production is difficult due to the lack of alternative livelihoods and/or the
23 affordability of other fuels (Smith et al. 2015; Zulu and Richardson 2013a; Jones et al. 2016a; World
24 Bank. 2009b).

25 **Use of agricultural residues and animal dung for bioenergy**

26 Although agricultural wastes and residues from almost any crop can be used in many cases for
27 bioenergy, excessive removal or reduction of forest (or agricultural) biomass can contribute to a loss
28 of soil carbon, which can also in turn contribute to land degradation (James et al. 2016; Blanco-
29 Canqui and Lal 2009a; Carvalho et al. 2016; Achat et al. 2015b; Stavi and Lal 2015). Removals are be
30 limited to levels at which problems of soil erosion, depletion of soil organic matter, soil nutrient
31 depletion and decline in crop yield are effectively mitigated (Ayamga et al. 2015a; Baudron et al.
32 2014; Blanco-Canqui and Lal 2009b). Application or recycling of residues may in some cases be
33 more valuable for soil improvement (*medium confidence*). Tao et al (2017) used leftover oil palm fruit
34 bunches and demonstrated that application of 30 to 90 t ha⁻¹ empty fruit bunches maintains high palm
35 oil yield with low temporal variability. A wide variety of wastes from palm oil harvesting can be used
36 for bioenergy, including annual crop residues (Go et al. 2019b; Ayamga et al. 2015b; Gardner et al.
37 2018b).

38 Animal dung is a low-quality fuel used where woody biomass is scarce, such as in South Asia and
39 some areas of eastern Africa (Duguma et al. 2014b; Behera et al. 2015b; Kumar Nath et al. 2013b).
40 Carbon and nutrient losses can be significant when animal dung is dried and burned as cake, whereas
41 using dung in a biodigester provides high-quality fuel and preserves nutrients in the by-product slurry
42 (Clemens et al. 2018; Gurung and Oh 2013b; Quinn et al. 2018b).

43 **Production and use of biochar**

44 Converting agricultural residues into biochar can also help to reverse trends of soil degradation (see
45 section 4.10.7). The positive effects of using biochar have been demonstrated in terms of soil
46 aggregate improvement, increase of exchangeable cations, cation exchange capacity, available P, soil
47 pH and carbon sequestration as well as increased crop yields (Huang et al. 2018; El-Naggar et al.

2018; Wang et al. 2018; Oladele et al. 2019; Blanco-Canqui and Lal 2009b). The level of biochar effectiveness varies depending on the kind of feedstock, soil properties and rate of application (Shaaban et al. 2018; Pokharel and Chang 2019). In addition to adding value to an energy product, the use of biochar offers a climate-smart approach to address agricultural productivity (Solomon and Lehmann 2017).

Relationship to food security and other SDGs

The population that is food insecure also intersects significantly with those relying heavily on traditional biomass such that poor and vulnerable populations often expend considerable time (gathering fuel) or use a significant share of household income for low quality energy services (Fuso Nerini et al. 2017; McCollum et al. 2018; Rao and Pachauri 2017; Pachauri et al. 2018; Muller and Yan 2018; Takama et al. 2012). Improvements in energy access and reduction or elimination of traditional biomass use thus have benefits across multiple SDGs (*medium evidence, high agreement*) (Masera et al. 2015; Rao and Pachauri 2017; Pachauri et al. 2018; Hoffmann et al. 2017; Jeuland et al. 2015; Takama et al. 2012; Gitau et al. 2019; Quinn et al. 2018b; Ruiz-Mercado and Masera 2015b; Duguma et al. 2014b; Sola et al. 2016b). Improved energy access contributes to adaptive capacity although charcoal production itself can also serve as a diversification or adaptation strategy (Perera et al. 2015; Ochieng et al. 2014; Sumiya 2016; Suckall et al. 2015; Jones et al. 2016b).

Socio-economic choices and shifts

When confronted with the limitations of higher-priced household energy alternatives, climate mitigation policies can result in trade-offs with health, energy access and other SDGs (Cameron et al. 2016; Fuso Nerini et al. 2018). The poorest households have no margin to pay for higher-cost efficient stoves; a focus on product-specific characteristics, user needs and/or making clean options more available would improve the market take-up (*medium confidence*) (Takama et al. 2012; Mudombi et al. 2018c; Khandelwal et al. 2017; Rosenthal et al. 2017; Cundale et al. 2017; Jürisoo et al. 2018). Subsidies for more efficient end-use technologies in combination with promotion of sustainable harvesting techniques would provide the highest emissions reductions while at the same time improving energy services (Cutz et al. 2017a).

Knowledge Gaps

Unlike analyses on modern energy sources, scientific assessments on traditional biomass use are complicated by its informal nature and the difficulty of tracing data and impacts; more systematic analytical efforts are needed to address this research gap (Cerutti et al. 2015). In general, traditional biomass use is associated with poverty. Therefore, efforts to reduce the dependence on fuelwood use are to be conducted in coherence with poverty alleviation (McCollum et al. 2018; Joana Specht et al. 2015b; Zulu and Richardson 2013b). The substantial potential co-benefits suggest that the traditional biomass sector remains under-researched and under-exploited in terms of cost-effective emissions reductions as well as for synergies between climate stabilisation goals and other SDGs.

7.6.3. Adaptive Climate Governance Responding to Uncertainty

In the 1990s, adaptive governance emerged from adaptive management (Holling 1978, 1986), combining resilience and complexity theory, and reflecting the trend of moving from government to governance (Hurlbert 2018b). Adaptive governance builds on multi-level and polycentric governance. Adaptive governance is “a process of resolving trade-offs and charting a course for sustainability” (Boyle, Michelle; Kay, James J.; Pond, 2001 at p. 28) through a range of “political, social, economic and administrative systems that develop, manage and distribute a resource in a manner promoting resilience through collaborative, flexible and learning based issue management across different

1 scales” (Margot A. Hurlbert, 2018 at p. 25). There is *medium evidence and medium agreement* that
2 few alternative governance theories handle processes of change characterised by nonlinear dynamics,
3 threshold effects, cascades and limited predictability; however, the majority of literature relates to the
4 United States or Canada (Karpouzoglou et al. 2016). Combining adaptive governance with other
5 theories has allowed good evaluation of important governance features such as power and politics,
6 inclusion and equity, short term and long term change, and the relationship between public policy and
7 adaptive governance (Karpouzoglou et al. 2016).

8 There is *robust evidence and high agreement* that resource and disaster crises are crises of governance
9 (Pahl-Wostl 2017b; Villagra and Quintana 2017; Gupta et al. 2013b). Adaptive governance of risk has
10 emerged in response to these crises and involves four critical pillars including 1) sustainability as a
11 response to environmental degradation, resource depletion and ecosystem service deterioration; 2)
12 recognition that governance is required as government is unable to resolve key societal and
13 environmental problems including climate change and complex problems; 3) mitigation is a means to
14 reduce vulnerability and avoid exposure; and 4) adaptation responds to changes in environmental
15 conditions (Fra.Paleo 2015).

16 Closely related to (and arguably components of) adaptive governance are adaptive management (see
17 7.5.4) (a regulatory environment that manages ecological system boundaries through hypothesis
18 testing, monitoring, and re-evaluation (Mostert et al. 2007)), adaptive co-management (flexible
19 community based resource management (Plummer and Baird 2013), and anticipatory governance
20 (flexible decision making through the use of scenario planning and reiterative policy review (Boyd et
21 al. 2015). Adaptive governance can be conceptualised as including multilevel governance with a
22 balance between top-down and bottom-up decision making that is performed by many actors
23 (including citizens) in both formal and informal networks, allowing policy measures and governance
24 arrangements to be tailored to local context and matched at the appropriate scale of the problem,
25 allowing for opportunities for experimentation and learning by individuals and social groups
26 (Rouillard et al. 2013; Hurlbert 2018b).

27 There is *high confidence* that anticipation is a key component of adaptive climate governance wherein
28 steering mechanisms in the present are developed to adapt to and/or shape uncertain futures (Vervoort
29 and Gupta 2018; Wiebe et al. 2018; Fuerth 2009). Effecting this anticipatory governance involves
30 simultaneously making short term decisions in the context of longer term policy visioning,
31 anticipating future climate change models and scenarios in order to realise a more sustainable future
32 (Bates and Saint-Pierre 2018; Serrao-Neumann et al. 2013; Boyd et al. 2015). Utilising the decision
33 making tools and practices in 7.5, policy makers operationalise anticipatory governance through a
34 foresight system considering future scenarios and models, a networked system for integrating this
35 knowledge into the policy process, a feedback system using indicators (see 7.5.5) to gauge
36 performance, an open-minded institutional culture allowing for hybrid and polycentric governance
37 (Fuerth and Faber 2013; Fuerth 2009).

38 There is *high confidence* that in order to manage uncertainty, natural resource governance systems
39 need to allow agencies and stakeholders to learn and change over time responding to ecosystem
40 changes and new information with different management strategies and practices that involve
41 experimentation (Camacho 2009; Young 2017b). There is an emerging literature on experimentation
42 in governance surrounding climate change and land use (Kivimaa et al. 2017a) including policies such
43 as REDD+ (Kaisa et al. 2017). Governance experiment literature could be in relation to scaling up
44 policies from the local level for greater application, or downscaling policies addressing broad
45 complex issues such as climate change, or addressing necessary change in social processes across
46 sectors (such as water energy and food) (Laakso et al. 2017). Successful development of new policy
47 instruments occurred in a governance experiment relating to coastal policy adapting to rising sea
48 levels and extreme weather events through planned retreat (Rocle and Salles 2018). Experiments in

1 emission trading between 1968 and 2000 in the United States of America helped to realise specific
2 models of governance and material practices through mutually supportive lab experiments and field
3 application that advanced collective knowledge (Voß and Simons 2018).

4 There is *high confidence* that a sustainable land management plan is dynamic and adaptive over time
5 to (unforeseen) future conditions by monitoring indicators as early warnings or signals of tipping
6 points initiating a process of change in policy pathway before a harmful threshold is reached
7 (Stephens et al. 2018, 2017; Haasnoot et al. 2013; Bloemen et al. 2018)(see 7.5.2.2). This process has
8 been applied in relation to coastal sea level rise starting with low risk, low cost measures and working
9 up to measures requiring greater investment after review and reevaluation (Barnett et al. 2014). A
10 first measure was stringent controls of new development, graduating to managed relocation of low
11 lying critical infrastructure, and eventually movement of habitable dwellings to more elevated parts
12 of town, as flooding and inundation triggers are experienced (Haasnoot et al. 2018; Lawrence et al.
13 2018; Barnett et al. 2014; Stephens et al. 2018). Nanda et al. (2018) apply the concept to a wetland in
14 Australia to identify a mix of short and long-term decisions, and Prober et al. (2017) develop
15 adaptation pathways for agricultural landscapes, also in Australia. Both studies identify that longer-
16 term decisions may involve a considerable change to institutional arrangements at different scales.
17 Viewing climate mitigation as a series of connected decisions over a long time period and not an
18 isolated decision, reduces the fragmentation and uncertainty endemic of models and effectiveness of
19 policy measures (Roelich and Gieseckam 2019).

20 There is *medium evidence and high agreement* that participatory processes in adaptive governance
21 within and across policy regimes overcome limitations of polycentric governance allowing priorities
22 to be set in sustainable development through rural land management and integrated water resource
23 management (Rouillard et al. 2013). Adaptive governance addresses large uncertainties and their
24 social amplification through differing perceptions of risk (Kasperson 2012; Fra.Paleo 2015) offering
25 an approach to co-evolve with risk by implementing policy mixes and assessing effectiveness in an
26 ongoing process, making mid-point corrections when necessary (Fra.Paleo 2015). In respect of
27 climate adaptation to coastal and riverine land erosion due to extreme weather events impacting
28 communities, adaptive governance offers the capacity to monitor local socio-economic processes and
29 implement dynamic locally informed institutional responses. In Alaska adaptive governance
30 responded to the dynamic risk of extreme weather events and issue of climate migration by providing
31 a continuum of policy from protection in place to community relocation, integrating across levels and
32 actors in a more effective and less costly response option than other governance systems (Bronen and
33 Chapin 2013). In comparison to other governance initiatives of ecosystem management aimed at
34 conservation and sustainable use of natural capital, adaptive governance has visible effects on natural
35 capital by monitoring, communicating and responding to ecosystem-wide changes at the landscape
36 level (Schultz et al. 2015). Adaptive governance can be applied to manage drought assistance as a
37 common property resource managing complex, interacting goals to create innovative policy options,
38 facilitated through nested and polycentric systems of governance effected by areas of natural resource
39 management including landscape care and watershed or catchment management groups (Nelson et al.
40 2008).

41 There is *medium evidence and high agreement* that transformational change is a necessary societal
42 response option to manage climate risks which is uniquely characterised by the depth of change
43 needed to reframe problems and change dominant mindsets, the scope of change needed (that is larger
44 than just a few people) and the speed of change required to reduce emissions (O' Brien et al. 2012;
45 Termeer et al. 2017). Transformation of governance occurs with changes in values to reflect an
46 understanding that the environmental crisis occurs in the context of our relation with the earth
47 (Hordijk et al. 2014; Pelling 2010). Transformation can happen by intervention strategies that enable
48 small in-depth wins, amplify these small wins through integration into existing practices, and unblock

1 stagnations (locked in structures) preventing transformation by confronting social and cognitive
2 fixations with counterintuitive interventions (Termeer et al. 2017). Iterative consideration of issues
3 and reformulation of policy instruments and response options facilitates transformation by allowing
4 experimentation (Monkelbaan 2019).

6 **Box 7.2 Adaptive Governance and interlinkages of food, fiber, water, energy and land**

7 Emerging literature and case studies recognise the connectedness of the environment and human
8 activities and the interrelationships of multiple resource-use practices in an attempt to understand
9 synergies and trade-offs (Albrecht et al. 2018). Sustainable adaptation - or actions contributing to
10 environmentally and socially sustainable development pathways (Eriksen et al. 2011) - requires
11 consideration of the interlinkage of different sectors (Rasul and Sharma 2016). Integrating
12 considerations can address sustainability (Hoff 2011) showing promise (Allan et al. 2015) for
13 effective adaptation to climate impacts in many drylands (Rasul and Sharma 2016).

14 Case studies of integrated water resources management (IWRM), landscape and ecosystem based
15 approaches illustrate important dimensions of institutions, institutional coordination, resource
16 coupling and local and global connections (Scott et al. 2011). Integrated governance, policy
17 coherence, and use of multi-functional systems are required to advance synergies across land, water,
18 energy and food sectors (Liu et al. 2017).

19 **Case Study: Flood and Food Security**

20 Between 2003–2013 floods were the most impacting natural disaster on crop production (FAO 2015b)
21 (albeit in certain contexts such riverine ecosystems and flood plain communities floods can be
22 beneficial).

23 In developing countries flood jeopardises primary access to food and impacts livelihoods. In
24 Bangladesh the 2007 flood reduced average consumption by 103Kcal/cap/day (worsening the existing
25 19.4% calories deficit) and in Pakistan the 2010 flood resulted in a loss of 205 Kcal/cap/day (or 8.5%
26 of the Pakistan average food supply). The Pakistan 2010 flood affected over 4.5 million workers, two
27 thirds employed in agriculture; 79% of farms lost greater than one half of their expected income
28 (Pacetti et al. 2017).

29 Policy instruments and response respond to the sequential and cascading impacts of flood. In a
30 Malawi study, flood impacts cascaded through labour, trade and transfer systems. First a harvest
31 failure occurred, followed by the decline of employment opportunities and reduction in real wages,
32 followed by a market failure or decline in trade, ultimately followed by a failure in informal safety
33 nets (Devereux 2007). Planned policy responses include those that address the sequential nature of
34 the cascading impacts starting with ‘productivity-enhancing safety nets’ addressing harvest failure,
35 then public works programmes addressing the decline in employment opportunities, followed by food
36 price subsidies to address the market failure, and finally food aid to address the failure of informal
37 safety nets (Devereux 2007). In another example in East Africa range lands, flood halted livestock
38 sales, food prices fell, and grain production ceased. Local food shortages couldn’t be supplemented
39 with imports due to destruction of transport links, and pastoral incomes were inadequate to purchase
40 food. Livestock diseases became rampant and eventually food shortages led to escalating prices. Due
41 to the contextual nature and timing of events, policy response initially addressed mobility and
42 resource access, and eventually longer term issues such as livestock disease (Little et al. 2001).

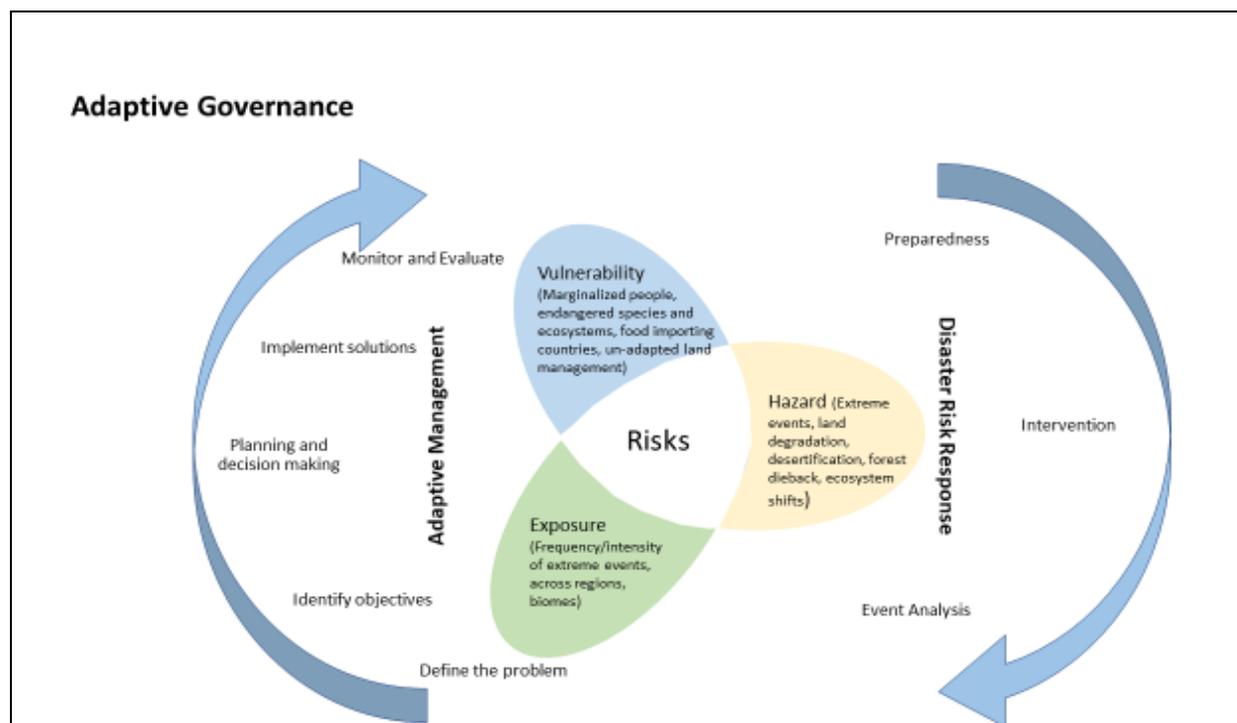
43 In North America floods are often described in terms of costs. For instance, the 1997 Red River
44 Basin flood cost Manitoba, Canada \$1 billion US and the United States of America, \$4 billion US in
45 terms of impact on agriculture and food production (Adaptation to Climate Change Team 2013). In
46 Canada floods accounted for 82% of disaster financial assistance spent from 2005–2014 (Public

1 Safety Canada 2017) and this cost may increase in the future. Future climate change may result in a
2 six foot rise in sea level by 2100 costing from USD 507 to 882 billion, affecting 300 American cities
3 (losing one half of their homes) and the wholesale loss of 36 cities (Lemann 2018).

4 Policy measures are important as an increasingly warming world may make post disaster assistance
5 and insurance increasingly unaffordable (Surminski et al. 2016). Historic legal mechanisms for
6 retreating from low lying and coastal areas have failed to encourage relocation of people out of flood
7 plains and areas of high risk (Stoa 2015). In some places cheap flood insurance and massive aid
8 programs have encouraged the populating of low-lying flood prone and coastal areas (Lemann 2018).
9 Although the state makes disaster assistance payments, it is local governments that determine
10 vulnerability through flood zone mapping, restrictions from building in flood zones, building
11 requirements (Stoa 2015), and integrated planning for flood. A comprehensive policy mix (see 7.4.8)
12 (implemented through adaptive management as illustrated on Figure 7.6) reduces vulnerability
13 (Hurlbert 2018b) (Hurlbert 2018a). Policy mixes that allow people to respond to disasters include
14 bankruptcy, insolvency rules, house protection from creditors, income minimums, and basic
15 agricultural implement protection laws. The portfolio of policies allows people to recover, and if
16 necessary migrate to other areas and occupations (Hurlbert 2018b).

17 At the international level, reactionary disaster response has evolved to proactive risk management that
18 combines adaptation and mitigation responses to ensure effective risk response, build resilient systems
19 and solve issues of structural social inequality (Innocenti and Albrito 2011). Advance measures of
20 preparedness are the main instruments to reduce fatalities and limit damage, as illustrated on the
21 figure below. The Sendai Declaration and Framework for Disaster Risk Reduction 2015-2030, is an
22 action plan to reduce mortality, the numbers of affected people and economic losses with four
23 priorities - understanding disaster risk, strengthening its governance to enhance the ability to manage
24 disaster risk, investing in resilience, and enhancing disaster preparedness. There is *medium evidence*
25 *and high agreement* that the Sendai Framework significantly refers to adaptive governance and could
26 be a window of opportunity to transform disaster risk reduction to address the causes of vulnerability
27 (Munene et al. 2018). Addressing disasters increasingly requires individual, household, community
28 and national planning and commitment to a new path of resilience and shared responsibility through
29 whole community engagement and linking private and public infrastructure interests (Rouillard et al.
30 2013). It is recommended that a vision and overarching framework of governance be adopted to allow
31 participation and coordination by government, nongovernmental organisations, researchers and the
32 private sector, individuals in the neighbourhood community. Disaster risk response is enhanced with
33 complementary structural and non-structural measures implemented together with measurable
34 scorecard indicators (Chen 2011).

1



2

3

Figure 7.8 Adaptive Governance

4 Adaptive management identifies and responds to exposure and vulnerability to land and climate
 5 change impacts by identifying problems and objectives, making decisions in relation to response
 6 options and instruments advancing response options in the context of uncertainty. These decisions are
 7 continuously monitored, evaluated and adjusted to changing conditions. Similarly disaster risk
 8 management responds to hazards through preparation, prevention, response, analysis, and
 9 reconstruction in an iterative process.

10

11 **7.6.4. Participation**

12 It is recognised that more benefits are derived when citizens actively participate in land and climate
 13 decision making, conservation, and policy formation (*high confidence*) (Jansujwicz et al. 2013)
 14 (Coenen and Coenen 2009; Hurlbert and Gupta 2015). Local leaders supported by strong laws,
 15 institutions, collaborative platforms, are able to draw on local knowledge, challenge external
 16 scientists, and find transparent and effective solutions for climate and land conflicts (Couvet and
 17 Prevot 2015; Johnson et al. 2017). Meaningful participation is more than providing
 18 technical/scientific information to citizens in order to accept decisions already made, but allows
 19 citizens to deliberate about climate change impacts to determine shared responsibilities creating
 20 genuine opportunity to construct, discuss, and promote alternatives (*high confidence*)(Lee et al. 2013;
 21 Armeni 2016; Pieraccini 2015)(Serrao-Neumann et al. 2015b; Armeni 2016). Participation is an
 22 emerging quality of collective-action and social-learning processes (see below) (Castella et al. 2014)
 23 when barriers for meaningful participation are surpassed (Clemens et al. 2015). The absence of
 24 systematic leadership, the lack of consensus on the place of direct citizen participation, and the limited
 25 scope and powers of participatory innovations limits the utility of participation (Fung 2015).

26 Multiple methods of participation exist, including multi-stakeholder forums, participatory scenario
 27 analyses, public forums and citizen juries (Coenen and Coenen 2009). No one method is superior, but
 28 each method must be tailored for local context (*high confidence*)(Blue and Medlock 2014; Voß and

1 Amelung 2016). Strategic innovation in developing policy initiatives requires a strategic adaptation
2 framework involving pluralistic and adaptive processes and use of boundary organisations (Head
3 2014).

4 The framing of a land and climate issue can influence the manner of public engagement (Hurlbert and
5 Gupta 2015) and studies have found local frames of climate change are particularly important
6 (Hornsey et al. 2016; Spence et al. 2012), emphasising diversity of perceptions to adaptation and
7 mitigation options (Capstick et al. 2015) (although Singh and Swanson (2017) found little evidence
8 framing impacted the perceived importance of climate change).

9 Recognition and use of indigenous and local knowledge (ILK) is an important element of
10 participatory approaches of various kinds. ILK can be used in decision-making on climate change
11 adaptation, Sustainable Land Management and food security at various scales and levels and is
12 important for long-term sustainability (*high confidence*). Cross-Chapter Box 13 discusses definitional
13 issues associated with ILK, evidence of its usefulness in responses to land-climate challenges,
14 constraints on its use, and possibilities for its incorporation in decision-making.
15

16 **Cross-Chapter Box 13: Indigenous and Local Knowledge**

17 John Morton (United Kingdom), Fatima Denton (The Gambia), James Ford (United Kingdom), Joyce
18 Kimutai (Kenya), Pamela McElwee (The United States of America), Marta Rivera Ferre (Spain),
19 Lindsay Stringer (United Kingdom)
20

21 Indigenous and local knowledge (ILK) can play a key role in climate change adaptation (*high*
22 *confidence*) (Mapfumo et al. 2017; Nyong et al. 2007b; Green and Raygorodetsky 2010; Speranza et
23 al. 2010; Alexander et al. 2011a; Leonard et al. 2013; Nakashima et al. 2013; Tschakert 2007). The
24 Summary for Policy-Makers of the Working Group II Contribution to the IPCC’s Fifth Assessment
25 Report (IPCC 2014b, p. 26) states that “Indigenous, local, and traditional knowledge systems and
26 practices, including indigenous peoples’ holistic view of community and environment, are a major
27 resource for adapting to climate change, but these have not been used consistently in existing
28 adaptation efforts. Integrating such forms of knowledge with existing practices increases the
29 effectiveness of adaptation” (see also Ford et al. 2016). The Special Report on Global Warming of
30 1.5 °C (IPCC 2018e; de Coninck et al. 2018) confirms the effectiveness and potential feasibility of
31 adaptation options based on ILK but also raises concerns that such knowledge systems are being
32 threatened by multiple socio-economic and environmental drivers (*high confidence*). The
33 Intergovernmental Platform on Biodiversity and Ecosystem Services (IPBES) Land Degradation and
34 Restoration Assessment (IPBES 2018) finds the same— that ILK can support adaptation to land
35 degradation but is threatened.
36

37 A variety of terminology has been used to describe indigenous and local knowledge: “Indigenous
38 knowledge”, “local knowledge”, “traditional knowledge”, “traditional ecological knowledge” and
39 other terms are used in overlapping and often inconsistent ways (Naess 2013). The Special Report on
40 Global Warming of 1.5°C (IPCC 2018a) reserves “indigenous knowledge” for culturally distinctive
41 ways of knowing associated with “societies with long histories of interaction with their natural
42 surroundings”, while using “local knowledge” for “understandings and skills developed by
43 individuals and populations, specific to the places where they live”, but not all research studies
44 observe this distinction. This Special Report generally uses “indigenous and local knowledge” (ILK)
45 as a combined term for these forms of knowledge, but in some sections the terminology used follows
46 that from the research literature assessed.
47

48 In contrast to scientific knowledge, ILK is context-specific, collective, transmitted informally, and is
49 multi-functional (Mistry and Berardi 2016; Naess 2013; Janif et al. 2016). Persson et al. (2018)
50 characterise ILK as “practical experience”, as locally-held knowledges are acquired through processes
51 of experience and interaction with the surrounding physical world. ILK is embedded in local

1 institutions (Naess 2013) and in cultural aspects of landscape and food systems (Fuller and Qingwen
2 2013; Koohafkan and Altieri 2011). ILK can encompass such diverse content as factual information
3 about the environment; guidance on rights and management; value statements about interactions with
4 others; and cosmologies and worldviews that influence how information is perceived and acted upon,
5 among other topics (Spoon 2014; Usher 2000).

6
7 This Cross-Chapter Box assesses evidence for the positive role of ILK in understanding climate
8 change and other environmental processes, and in managing land sustainably in the face of climate
9 change, desertification, land degradation and food insecurity. It also assesses constraints on and
10 threats to the use of ILK in these challenges, and processes by which ILK can be incorporated in
11 decision-making and governance processes.

12 *ILK in understanding and responding to climate change impacts*

13
14
15 ILK can play a role in understanding climate change and other environmental processes, particularly
16 where formal data collection is sparse (Alexander et al. 2011a; Schick et al. 2018), and can contribute
17 to accurate predictions of impending environmental change (Green and Raygorodetsky 2010; Orlove
18 et al. 2010) (medium confidence). Both at global level (Alexander et al. 2011a; Green and
19 Raygorodetsky 2010), and local level (Speranza et al. 2010; Ayanlade et al. 2017), strong correlations
20 between local perceptions of climate change and meteorological data have been shown, as calendars,
21 almanacs, and other seasonal and interannual systems knowledge embedded in ILK hold information
22 about environmental baselines (Orlove et al. 2010; Cochran et al. 2016).

23
24 ILK is strongly associated with sustainable management of natural resources, including land, and with
25 autonomous adaptation to climate variability and change, while also serving as a resource for
26 externally-facilitated adaptation (Stringer et al. 2009). For example, women’s traditional knowledge
27 adds value to a society’s knowledge base and supports climate change adaptation practices (Lane and
28 McNaught 2009). In dryland environments, populations have historically demonstrated remarkable
29 resilience and innovation to cope with high climatic variability, manage dynamic interactions between
30 local communities and ecosystems, and sustain livelihoods (Safriel and Adeel 2008; Davies 2017).
31 There is high confidence that pastoralists have created formal and informal institutions based on ILK
32 for regulating grazing, collection and cutting of herbs and wood, and use of forests across the Middle
33 East and North Africa (Louhaichi and Tastad 2010; Domínguez 2014; Auclair et al. 2011), Mongolia
34 (Fernandez-Gimenez 2000), The Horn of Africa (Oba 2013) and the Sahel (Krätli and Schareika
35 2010). Herders in both the Horn of Africa and the Sahel have developed complex livestock breeding
36 and selection systems for their dryland environment (Krätli 2008; Fre 2018). Numerous traditional
37 water harvesting techniques are used across the drylands to adapt to climate variability: planting pits
38 (“zai”, “ngoro”) and micro-basins and contouring hill slopes and terracing (Biazin et al. 2012),
39 alongside the traditional “ndiva” water harvesting system in Tanzania to capture runoff in community-
40 managed micro-dams for small-scale irrigation (Enfors and Gordon 2008).

41
42 Across diverse agro-ecological systems, ILK is the basis for traditional practices to manage the
43 landscape and sustain food production, while delivering co-benefits in the form of biodiversity and
44 ecosystem resilience at a landscape scale (high confidence). Flexibility and adaptiveness are
45 hallmarks of such systems (Richards 1985; Biggs et al. 2013), and documented examples include:
46 traditional integrated watershed management in the Philippines (Camacho et al. 2016); widespread
47 use of terracing with benefits in cases of both intensifying and decreasing rainfall (Arnáez et al. 2015;
48 Chen et al. 2017b) and management of water harvesting and local irrigation systems in the Indo-
49 Gangetic Plain (Rivera-Ferre et al. 2016). Rice cultivation in East Borneo is sustained by traditional
50 forms of shifting cultivation, often involving intercropping of rice with bananas, cassava and other
51 food crops (Siahaya et al. 2016), although the use of fire in land clearance implies trade-offs for
52 climate change mitigation which have been sparsely assessed. Indigenous practices for enhanced soil
53 fertility have been documented among South Asian farmers (Chandra et al. 2011; Dey and Sarkar
54 2011) and among Mayan farmers where management of carbon has positive impacts on mitigation
55 (Falkowski et al. 2016). Korean traditional groves or “bibosoop” have been shown to reduce wind

1 speed and evaporation in agricultural landscapes (Koh et al. 2010). Particularly in the context of
2 changing climates, agriculture based on ILK that focuses on biodiversification, soil management, and
3 sustainable water harvesting holds promise for long-term resilience (Altieri and Nicholls 2017) and
4 rehabilitation of degraded land (Maikhuri et al. 1997). ILK is also important in other forms of
5 ecosystem management, such as forests and wetlands, which may be conserved by efforts such as
6 sacred sites (Ens et al. 2015; Pungetti et al. 2012) and ILK can play an important role in ecological
7 restoration efforts, including for carbon sinks, through knowledge surrounding species selection and
8 understanding of ecosystem processes, like fire (Kimmerer 2000).

9 10 *Constraints on the use of ILK*

11
12 Use of ILK as a resource in responding to climate change can be constrained in at least three ways
13 (high confidence). Firstly the rate of climate change and the scale of its impacts may render
14 incremental adaptation based on the ILK of smallholders and others, less relevant and less effective
15 (Lane and McNaught 2009; Orłowsky and Seneviratne 2012; Huang et al. 2016; Morton 2017).
16 Secondly, maintenance and transmission of ILK across generations may be disrupted by e.g.: formal
17 education, missionary activity, livelihood diversification away from agriculture, and a general
18 perception that ILK is outdated and unfavourably contrasted with scientific knowledge (Speranza et
19 al. 2010), and by HIV-related mortality (White and Morton 2005). Urbanisation can erode ILK,
20 although ILK is constantly evolving, and becoming integrated into urban environments (Júnior et al.
21 2016; Oteros-Rozas et al. 2013; van Andel and Carvalheiro 2013). Thirdly, ILK holders are
22 experiencing difficulty in using ILK due to loss of access to resources, such as through large-scale
23 land acquisition (Siahaya et al. 2016; Speranza et al. 2010; de Coninck et al. 2018) and the increasing
24 globalisation of food systems and integration into global market economy also threatens to erode ILK
25 (Gómez-Baggethun et al. 2010; Oteros-Rozas et al. 2013; McCarter et al. 2014). The potential role
26 that ILK can play in adaptation at the local level depends on the configuration of a policy-institutions-
27 knowledge nexus (Stringer et al. 2018), which includes power relations across levels and interactions
28 with government strategies (Alexander et al. 2011b; Naess 2013).

29 30 *Incorporation of ILK in decision-making*

31
32 ILK can be used in decision-making on climate change adaptation, Sustainable Land Management
33 and food security at various scales and levels and is important for long-term sustainability (high
34 confidence). Respect for ILK is both a requirement and an entry strategy for participatory climate
35 action planning and effective communication of climate action strategies (Nyong et al. 2007b). The
36 nature, source, and mode of knowledge generation are critical to ensure that sustainable solutions are
37 community-owned and fully integrated within the local context (Mistry and Berardi 2016). Integrating
38 ILK with scientific information is a prerequisite for such community-owned solutions. Scientists can
39 engage farmers as experts in processes of knowledge co-production (Oliver et al. 2012), helping to
40 introduce, implement, adapt and promote locally appropriate responses (Schwilch et al. 2011).
41 Specific approaches to decision-making that aim to integrate indigenous and local knowledge include
42 some versions of decision support systems (Jones et al. 2014) as well as citizen science and
43 participatory modelling (Tengö et al. 2014).

44
45 ILK can be deployed in the practice of climate governance especially at the local level where actions
46 are informed by the principles of decentralisation and autonomy (Chanza and de Wit 2016;
47 Harmsworth and Awatere 2013). International environmental agreements also are increasingly
48 including attention to ILK and diverse cultural perspectives, for reasons of social justice and inclusive
49 decision-making (Brondizio and Tourneau 2016). However, the context-specific, and dynamic nature
50 of ILK and its embeddedness in local institutions and power relations needs consideration (Naess
51 2013). It is also important to take a gendered approach so as not to further marginalise certain
52 knowledge, as men and women hold different knowledge, expertise and transmission patterns (Díaz-
53 Reviriego et al. 2017).

1 **Citizen Science**

2 Citizen science is a democratic approach to science involving citizens in collecting, classifying, and
3 interpreting data to influence policy and assist decision processes, including issues relevant to the
4 environment (Kullenberg and Kasperowski 2016). It has flourished in recent years due to easily
5 available technical tools for collecting and disseminating information (e.g., cell phone-based apps,
6 cloud-based services, ground sensors, drone imagery, and others), recognition of its free source of
7 labour, and requirements of funding agencies for project related outreach (Silvertown 2009). There is
8 significant potential for combining citizen science and participatory modelling to obtain favourable
9 outcomes and improve environmental decision making (*medium confidence*) (Gray et al. 2017).
10 Citizen participation in land use simulation integrates stakeholders' preferences through the
11 generation of parameters in analytical and discursive approaches (Hewitt et al. 2014), and thereby
12 supports the translation of narrative scenarios to quantitative outputs (Mallampalli et al. 2016),
13 supports the development of digital tools to be used in co-designing decision making participatory
14 structures (Bommel et al. 2014), and supports the use of games to understand the preferences of local
15 decision making when exploring various balanced policies about risks (Adam et al. 2016).

16 There is *medium confidence* that citizen science improves sustainable land management through
17 mediating and facilitating landscape conservation decision making and planning, as well as boosting
18 environmental awareness and advocacy (Lange and Hehl-Lange 2011; Bonsu et al. 2017; Graham et
19 al. 2015) (Bonsu et al. 2017) (Lange and Hehl-Lange 2011) (Sayer, J. Margules, C., Boedhihartono
20 2015) (McKinley et al. 2017) (Johnson et al. 2017, 2014) (Gray et al. 2017). One study found limited
21 evidence of direct conservation impact (Ballard et al. 2017) and most of the cases derive from rich
22 industrialised countries (Loos et al. 2015). There are many practical challenges to the concept of
23 citizen science at the local level, which include differing methods and the lack of universal
24 implementation framework (Conrad and Hilchey 2011; Jalbert and Kinchy 2016; Stone et al. 2014).
25 Uncertainty related to citizen science needs to be recognised and managed (Swanson et al. 2016; Bird
26 et al. 2014; Lin et al. 2015) and citizen science projects around the world need better coordination to
27 understand significant issues, such as climate change (Bonney et al. 2014).
28

29 **Participation, Collective Action, and Social Learning**

30 As land and climate issues cannot be solved by one individual, a diverse collective action issue exists
31 for land use policies and planning practices (Moroni 2018) at local, national, and regional levels.
32 Collective action involves individuals and communities in land planning processes in order to
33 determine successful climate adaptation and mitigation (Nkoana et al. 2017) (Liu and Ravenscroft
34 2017) (Nieto-Romero et al. 2016; Nikolakis et al. 2016), or as Sarzynski (2015) finds, a community
35 'pulling together' to solve common adaptation and land planning issues.

36 Collective action offers solutions for emerging land and climate change risks, including strategies
37 that target maintenance or change of land use practices, increase livelihood security, risk share
38 through pooling, and sometimes also aim to promote social and economic goals such as reducing
39 poverty (Samaddar et al. 2015)(Andersson and Gabrielsson 2012). Collective action has resulted in
40 the successful implementation of national-level land transfer policies (Liu and Ravenscroft 2017),
41 rural development and land sparing (Jelsma et al. 2017), and the development of tools to identify
42 shared objectives, trade-offs and barriers to land management (Nieto-Romero et al. 2016; Nikolakis
43 et al. 2016). Collective action can also produce mutually binding agreements, government regulation,
44 privatisation, and incentive systems (IPCC 2014c).

45 Successful collective action requires understanding and implementation of factors that determine
46 successful participation in climate adaptation and mitigation (Nkoana et al. 2017). These include
47 ownership, empowerment or self-reliance, time effectiveness, economic and behavioural interests,
48 livelihood security, and the requirement for plan implementation (Samaddar et al. 2015; Djurfeldt
49 et al. 2018) (Sánchez and Maseda 2016). In a UK study, dynamic trust relations among members
50 around specific issues, determined the potential of agri-environmental schemes to offer landscape-

1 scale environmental protection (Riley et al. 2018). Collective action is context specific and rarely
2 scaled up or replicated in other places (Samaddar et al. 2015).

3 Collective action in land use policy has been shown to be more effective when implemented as
4 bundles of actions rather than as single-issue actions. For example, land tenure, food security, and
5 market access can mutually reinforce each other when they are interconnected (Corsi et al. 2017). For
6 example, (Liu and Ravenscroft 2017) found that financial incentives embedded in collective forest
7 reforms in China have increased forest land and labour inputs in forestry.

8 A product of participation, equally important in practical terms, is social learning (*high confidence*)
9 (Reed et al. 2010) (Dryzek and Pickering 2017) (Gupta 2014), which is learning in and with social
10 groups through interaction (Argyris 1999) including collaboration and organisation which occurs in
11 networks of interdependent stakeholders (Mostert et al. 2007). Social learning is defined as a change
12 in understanding measured by a change in behaviour, and perhaps worldview, by individuals and
13 wider social units, communities of practice and social networks (Reed et al. 2010) (Gupta 2014).
14 Social learning is an important factor contributing to long-term climate adaptation whereby
15 individuals and organisations engage in a multi-step social process, managing different framings of
16 issues while raising awareness of climate and land risks and opportunities, exploring policy options
17 and institutionalising new rights, responsibilities, feedback and learning processes (Tàbara et al.
18 2010). It is important for engaging with uncertainty (Newig et al. 2010) and addressing the increasing
19 unequal geography of food security (Sonnino et al. 2014).

20 Social learning is achieved through reflexivity or the ability of a social structure, process, or set of
21 ideas to reconfigure itself after reflection on performance though open-minded people interacting
22 iteratively to produce reasonable and well-informed opinions (Dryzek and Pickering 2017). These
23 processes develop through skilled facilitation attending to social difference and power resulting in a
24 shared view of how change might happen (Harvey et al. 2012; Ensor and Harvey 2015). When
25 combined with collective action, social learning can make transformative change measured by a
26 change in worldviews (beliefs about the world and reality) and understanding of power dynamics
27 (Gupta 2014) (Bamberg et al. 2015).

28 **7.6.5. Land Tenure**

29 Land tenure, defined as “the terms under which land and natural resources are held by individuals,
30 households or social groups”, is a key dimension in any discussion of land-climate interactions,
31 including the prospects for both adaptation and land-based mitigation, and possible impacts on tenure
32 and thus land security of both climate change and climate action (Quan and Dyer 2008) (*medium*
33 *evidence, high agreement*).

34 Discussion of land tenure in the context of land-climate interactions in developing countries needs to
35 consider the prevalence of informal, customary and modified customary systems of land tenure:
36 estimates range widely, but perhaps as much as 65% of the world’s total land area is managed under
37 some form of these local, customary or communal tenure systems, and only a small fraction of this
38 (around 15%) is formally recognised by governments (Rights and Resources Initiative 2015a). These
39 customary land rights can extend across many categories of land, but are difficult to assess properly
40 due to poor reporting, lack of legal recognition, and lack of access to reporting systems by indigenous
41 and rural peoples (Rights and Resources Initiative 2018a). Around 521 million ha of forest land is
42 estimated to be legally owned, recognised, or designated for use by indigenous and local communities
43 as of 2017 (Rights and Resources Initiative 2018b), predominantly in Latin America, followed by
44 Asia. However in India approximately 40 million ha of forest land is managed under customary rights
45 not recognised by the government (Rights and Resources Initiative 2015b). In 2005 only 1% of land
46 in Africa was legally registered (Easterly 2008a).

47 Much of the world's carbon is stored in the biomass and soil on the territories of customary
48 landowners including indigenous peoples (Walker et al. 2014; Garnett et al. 2018), making securing

1 of these land tenure regimes vital in land and climate protection. These lands are estimated to hold at
2 least 293 GtC of carbon, of which around one-third (72 GtC) is located in areas where indigenous
3 peoples and local communities lack formal recognition of their tenure rights (Frechette et al. 2018).

4 Understanding the interactions between land tenure and climate change has to be based on underlying
5 understanding of land tenure and land policy and how they relate to sustainable development,
6 especially in low- and middle-income countries: such understandings have changed considerably over
7 the last three decades, and now show that informal or customary systems can provide secure tenure
8 (Toulmin and Quan 2000). For smallholder systems, (Bruce and Migot-Adholla 1994) among other
9 authors established that African customary tenure can provide the necessary security for long-term
10 investments in farm fertility such as tree-planting. For pastoral systems, (Behnke 1994; Lane and
11 Moorehead 1995) and other authors showed the rationality of communal tenure in situations of
12 environmental variability and herd mobility. However, where customary systems are unrecognised or
13 weakened by governments or the rights from them undocumented or unenforced, tenure insecurity
14 may result (Lane 1998; Toulmin and Quan 2000). There is strong empirical evidence of the links
15 between secure communal tenure and lower deforestation rates, particularly in intact forests
16 (Nepstad et al., 2006; Persha, Agrawal, & Chhatre, 2011; Vergara-Asenjo & Potvin, 2014). Securing
17 and recognising tenure for indigenous communities (such as through revisions to legal or policy
18 frameworks) has been shown to be highly cost effective in reducing deforestation and improving land
19 management in certain contexts, and is therefore also apt to help improve indigenous communities'
20 ability to adapt to climate changes (Suzuki 2012; Balooni et al. 2008; Ceddia et al. 2015; Pacheco et
21 al. 2012; Holland et al. 2017).

22 Rights to water for agriculture or livestock are linked to land tenure in complex ways still little
23 understood and neglected by policy-makers and planners (Cotula 2006a). Provision of water
24 infrastructure tends to increase land values, but irrigation schemes often entail reallocation of land
25 rights (Cotula 2006b) and new inequalities based on water availability such as the creation of a
26 category of tailenders in large-scale irrigation (Chambers 1988) and disruption of pastoral grazing
27 patterns through use of riverine land (Behnke and Kerven 2013).

28 Understanding of land tenure under climate change also has to take account of the growth in large-
29 scale land acquisitions (LSLAs), also referred to as land-grabbing, in developing countries. These
30 LSLAs are defined by acquisition of more than 200 ha per deal (Messerli et al. 2014a). Klaus
31 Deininger (2011) links the growth in demand for land to the 2007-2008 food price spike, and
32 demonstrates that high levels of demand for land at the country level are statistically associated with
33 weak recognition of land rights. Land grabs, where LSLAs occur despite local use of lands, are often
34 driven by direct collaboration of politicians, government officials and land agencies (Koechlin et al.
35 2016), involving corruption of governmental land agencies, failures to register community land claims
36 and illegal lands uses and lack of the rule of law and enforcement in resource extraction frontiers
37 (Borras Jr et al. 2011). Though data is poor, overall, small and medium scale domestic investment has
38 in fact been more important than foreign investment (Deininger 2011; Cotula et al. 2014). There are
39 variations in estimates of the scale of large-scale land acquisitions: the Nolte et al. (2016) report
40 concluded deals totalling 42.2 million ha worldwide. Cotula et al. (2014) using cross-checked data for
41 completed lease agreements in Ethiopia, Ghana and Tanzania conclude they cover 1.9%, 1.9% and
42 1.1% respectively of each country's total land suitable for agriculture. The literature expresses
43 different views on whether these acquisitions concern marginal lands or lands already in use thereby
44 displacing existing users (Messerli et al. 2014b). Land-grabbing is associated with and may be
45 motivated by the acquisition of rights to water, and erosion of those rights for other users such as
46 those downstream (Mehta et al. 2012). Quantification of the acquisition of water rights resulting from
47 LSLAs raises major issues of definition, data availability, and measurement. One estimate of the total
48 acquisition of gross irrigation water associated with land-grabbing across the 24 countries most
49 affected is 280 billion m³ (Rulli et al. 2013).

50 While some authors see LSLAs as investments that can contribute to more efficient food production at
51 larger scales (World Bank 2011; Deininger and Byerlee 2012), others have warned that local food
52 security may be threatened by them (Daniel 2011; Golay and Biglino 2013; Lavers 2012). Reports
53 suggest that recent land grabbing has affected 12 million people globally in terms of declines in

1 welfare (Adnan 2013; Davis et al. 2014). De Schutter (2011) argues that large-scale land acquisitions
2 will a) result in types of farming less liable to reduce poverty than smallholder systems, b) increase
3 local vulnerability to food price shocks by favouring export agriculture and c) accelerate the
4 development of a market for land with detrimental impacts on smallholders and those depending on
5 common property resources. Land grabbing can threaten not only agricultural lands of farmers, but
6 also protected ecosystems, like forests and wetlands (Hunsberger et al. 2017; Carter et al. 2017; Ehara
7 et al. 2018).

8 The primary mechanisms for combatting LSLAs have included restrictions on the size of land sales
9 (Fairbairn 2015); pressure on agribusiness companies to agree to the Voluntary Guidelines on the
10 Responsible Governance of Tenure of Land, Fisheries and Forests in the Context of National Food
11 Security, known as the VGGT, or similar principles (Collins 2014; Goetz 2013); attempts to repeal
12 biofuels standards (Palmer 2014); strengthening of existing land law and land registration systems
13 (Bebbington et al. 2018); use of community monitoring systems (Sheil et al. 2015); and direct protests
14 against the land acquisitions (Hall et al. 2015; Fameree 2016).

15 Table 7.7 sets out, in highly summarised form, some key findings on the multi-directional inter-
16 relations between land tenure and climate change, with particular reference to developing countries.
17 The rows represent different categories of landscape or resource systems. For each system the second
18 column summarises current understandings on land tenure and sustainable development, in many case
19 predating concerns over climate change. The third column summarises the most important
20 implications of land tenure systems, policy about land tenure, and the implementation of that policy,
21 for vulnerability and adaptation to climate change, and the fourth gives a similar summary for
22 mitigation of climate change. The fifth column summarises key findings on how climate change and
23 climate action (both adaptation and mitigation) will impact land tenure, and the final column, findings
24 on implications of climate change for evolving land policy.

25

1

Table 7.7 Major Findings on the Interactions between Land Tenure and Climate Change

| Landscape or natural resource system | State of understanding of land tenure, land policy and sustainable development | Implications of land tenure for vulnerability and adaptation to climate change | Implications of land tenure for mitigation of climate change | Impacts of climate change and climate action on land tenure | Implications of climate change and climate action for land policy |
|--------------------------------------|---|---|---|--|--|
| Smallholder cropland | In South Asia and Latin America the poor suffer from limited access including insecure tenancies, though this has been partially alleviated by land reform. ¹ In Africa informal/customary systems may provide considerable land tenure security and enable long-term investment in land management, but are increasingly weakened by demographic pressures on available land resources increase. However, creation of freehold rights through conventional land titling is not a necessary condition for tenure security and may be cost-ineffective or counter-productive. ^{2,3,4,5} Alternative approaches utilising low cost technologies and participatory methods are available. ⁶ Secure and defensible land tenure, including modified customary tenure, has been positively correlated with food production increases. ^{7,8,9} | Insecure land rights are one factor deterring adaptation and accentuating vulnerability. ^{10,11} Specific dimensions of inequity in customary systems may act as constraints on adaptation in different contexts. ¹² LSLAs may be associated with monoculture and other unsustainable land use practices, have negative consequences for soil degradation ¹³ and disincentivise more sustainable forms of agriculture. ¹⁴ | Secure land rights, including through customary systems, can incentivise farmers to adopt long-term climate-smart practices, ¹⁵ e.g., planting trees in mixed cropland/forest systems. ¹⁶ | Increased frequency and intensity of extreme weather can lead to displacement and effective loss of land rights. ¹⁷ REDD+ programmes tend slightly to increase land tenure insecurity on agricultural forest frontier lands, - but not in forests. ¹⁸ | Landscape governance and resource tenure reforms at farm and community levels can facilitate and incentivise planning for landscape management and enable the integration of adaptation and mitigation strategies. ¹¹ |
| Rangelands | Communal management of rangelands in pastoral systems is a rational and internally sustainable response to climate variability and the need for mobility. Policies favouring individual or small group land-tenure may have negative impacts on both ecosystems and livelihoods. ^{19,20,21} | Many pastoralists in lands at risk from desertification do not have secure land tenure, and erosion of traditional communal rangeland tenure has been identified as a determinant of increasing vulnerability to drought and climate change and as a driver of dryland | Where pastoralists' traditional land use does not have legal recognition, or where pastoralists are unable to exclude others from land use, this presents significant challenges for carbon sequestration initiatives. ^{27,28} | Increasing conflict on rangelands is a possible result of climate change and environmental pressures, but depends on local institutions. ²⁹ Where land use rights for pastoralists are absent or unenforced, demonstrated potential for carbon sequestration may assist | Carbon sequestration initiatives on rangelands may require clarification and maintenance of land rights. ^{27,28} |

| | | degradation. ^{22,23,24,25,26} | | advocacy. ²⁸ | |
|-------------------------------------|--|---|---|---|---|
| Forests | Poor management of state and open-access forests has been combatted in recent years by a move towards forest decentralisation and community co-management. ^{30,31,32,33,34,35} Land tenure systems have complex interactions with deforestation processes. Land tenure security is generally associated with less deforestation, regardless of whether the tenure form is private, customary or communal. ^{33,36,37,38} Historical injustices towards forest dwellers can be ameliorated with appropriate policy, e.g., 2006 Forest Rights Act in India. ³⁹ | Land tenure security can lead to improved adaptation outcomes ^{40,41,42,43} but land tenure policy for forests that focuses narrowly on cultivation has limited ability to reduce ecological vulnerability or enhance adaptation. ³⁹ Secure rights to land and forest resources can facilitate efforts to stabilise shifting cultivation and promote more sustainable resource use if appropriate technical and market support are available. ⁴⁴ | Land tenure insecurity has been identified as a key driver of deforestation and land degradation leading to loss of sinks and creating sources of GHGs ^{45,46,47,48,49} . While land tenure systems interact with land-based mitigation actions in complex ways, ³⁶ forest decentralisation and community co-management has shown considerable success in slowing forest loss and contributing to carbon mitigation. ^{30,31,32,33,34,35} Communal tenure systems may lower transaction costs for REDD+ schemes, though with risk of elite capture of payments. ¹⁶ | Findings on both direction of change in tenure security and extent to which this has been influenced by REDD+ are very diverse. ^m The implications of land-based mitigation (e.g., BECCS) on land tenure systems is currently understudied, but evidence from biofuels expansion shows negative impacts on local livelihoods and loss of forest sinks where LSLAs override local land tenure. ^{50,51} | Forest tenure policies under climate change need to accommodate and enable evolving and shifting boundaries linked to changing forest livelihoods. ¹⁰ REDD+ programmes need to be integrated with national-level forest tenure reform. ¹⁸ |
| Poor and informal urban settlements | Residents of poor and informal urban settlements enjoy varying degrees of tenure security from different forms of tenure. Security will be increased by building on de facto rights rather than through abrupt changes in tenure systems. ⁵² | Public land on the outskirts of urban areas can be used to adapt to increasing flood risks by protecting natural assets. ⁵³ Secure land titles in hazardous locations may make occupants reluctant to move and raise the costs of compensation and resettlement. ¹⁷ | Urban land use strategies such as tree planting, establishing public parks, can save energy usage by moderating urban temperature and protect human settlement from natural disaster such as flooding or heatwaves. ⁵⁴ | Without proper planning, climate hazards can undermine efforts to recognise and strengthen informal tenure rights without proper planning. ^{55,56} | Climate risks increase the requirements for land use planning and settlement that increases tenure security, with direct involvement of residents, improved use of public land, and innovative collaboration with private and traditional land owners. ^{56,57} |
| Riverscapes and riparian fringes | Well-defined but spatially flexible community tenure can support regulated and sustainable artisanal capture | Unequal land rights and absence of land management | Mitigation measures such as protection of riparian forests and grasslands can | | Secured but spatially flexible tenure will enable climate change mitigation |

| | | | | | |
|--|---|---|---|--|--|
| | fisheries and biodiversity. ^{58,59,60,61,62,63,64} | arrangements in floodplains increases vulnerability and constrains adaptation. ⁶⁵ Marginalised or landless fisherfolk will be empowered by tenurial rights and associated identity to respond more effectively to ecological changes in riverscapes including riparian zones. ^{66,67,68,69} | potentially play a major role, provided rights to land and trees are sufficiently clear. ^{70,71} | | in riverscapes to be synergised with local livelihoods and ecological security. ^{67,72} |
|--|---|---|---|--|--|

1 Sources: 1) Binswanger et al. 1995 2) Schlager and Ostrom 1992 3) Toulmin and Quan 2000 4) Bruce and Migot-Adholla 1994 5) Easterly 2008 6) McCall and Dunn 2012 7) Maxwell and
 2 Wiebe 1999 8) Holden and Ghebru 2016 9) Corsi et al. 2017 10) Quan et al. 2017 11) Harvey et al. 2014 12) Antwi-Agyei et al. 2015 13) Balehegn, 2015 14) Friis & Nielsen, 2016 15)
 3 Scherr et al. 2012 16) Barbier and Tesfaw 2012 17) Mitchell 2010 18) Sunderlin et al. 2018 19) Behnke 1994 20) Lane and Moorehead 1995 21) Davies et al. 2015 22) Morton 2007 23)
 4 López-i-Gelats et al. 2016 24) Oba 1994 25) Fraser et al. 2011 26) Dougill et al. 2011 27) Roncoli et al. 2007. 28) Tennigkeit and Wilkes 2008 29) Adano et al. 2012 30) Agrawal, Chhatre,
 5 & Hardin, 2008 31) Chhatre & Agrawal, 2009 32) Gabay & Alam, 2017 33) Holland et al., 2017 34) Larson & Pulhin, 2012 35) Pagdee, Kim, & Daugherty, 2006) 36) Robinson et al. 2014
 6 37) Blackman et al. 2017 38) Nelson et al. 2001; 38) Ramnath 2008 40) Suzuki 2012 41) Balooni et al. 2008 42) Ceddia et al. 2015 43) Pacheco et al. 2012) 44) Garnett et al. 2013 45)
 7 Clover & Eriksen, 2009 46) Damnyag, Saastamoinen, Appiah, & Pappinen, 2012 47) Finley-Brook, 2007 48) Robinson, Holland, & Naughton-Treves, 2014 49) Stickler, Huntington, Haflett,
 8 Petrova, & Bouvier, 2017 50) Romijn, 2011 51) Aha & Ayitey, 2017 52) Payne 2001 53) Barbedo et al. 2015 54) Zhao et al. 2018 55) Satterthwaite et al. 2018 56) Mitchell et al. 2015 57)
 9 Satterthwaite 2007 58) Thomas 1996 59) Welcomme et al. 2010 60) Silvano and Valbo-Jørgensen 2008 61) Biermann et al. 2012 62) Abbott et al. 2007 63) Béné et al. 2011 64) McGrath
 10 et al. 1993 65) Barkat et al. 2001 66) FAO 2015 67) Hall et al. 2013 68) Berkes 2001 69) ISO 2017 70) Rocheleau and Edmunds 1997 71) Baird and Dearden 2003 72) Béné et al. 2010.

11

12

1 In drylands, weak land tenure security, either for households disadvantaged within a customary tenure
2 system or more widely as such a system is eroded, can be associated with increased vulnerability and
3 decreased adaptive capacity (*limited evidence, high agreement*). There is *medium evidence* and
4 *medium agreement* that land titling and recognition programs, particularly those that authorise and
5 respect indigenous and communal tenure, can lead to improved management of forests, including for
6 carbon storage (Suzuki 2012; Balooni et al. 2008; Ceddia et al. 2015; Pacheco et al. 2012), primarily
7 by providing legally secure mechanisms for exclusion of others (Nelson et al. 2001; Blackman et al.
8 2017). However, these titling programs are highly context-dependent and there is also evidence that
9 titling can exclude community and common management, leading to more confusion over land rights,
10 not less, where poorly implemented (Broegaard et al. 2017). For all the systems, an important finding
11 is that land policies can provide both security and flexibility in the face of climate change, but through
12 a diversity of forms and approaches (recognition of customary tenure, community mapping,
13 redistribution, decentralisation, co-management, regulation of rental markets, strengthening the
14 negotiating position of the poor) rather than sole focus on freehold title (Quan & Dyer, 2008; K
15 Deininger & Feder, 2009; St. Martin, 2009) (*medium evidence, high agreement*). Land policy can be
16 climate-proofed and integrated with national policies such as NAPAs (Quan and Dyer 2008). Land
17 administration systems have a vital role in providing land tenure security, especially for the poor,
18 especially when linked to an expanded range of information relevant to mitigation and adaptation
19 (Quan and Dyer 2008; van der Molen and Mitchell 2016). Challenges to such a role include outdated
20 and overlapping national land and forest tenure laws, which often fail to recognise community
21 property rights and corruption in land administration (Monterrosso et al. 2017), as well as lack of
22 political will and the costs of improving land administration programs (Deininger and Feder 2009).

23

24 **7.6.6. Institutional dimensions of adaptive governance**

25 Institutional systems that demonstrate the institutional dimensions, or indicators, in Table 7.8 enhance
26 the adaptive capacity of the socio-ecological system to a greater degree than institutional systems that
27 do not demonstrate these dimensions (*high confidence*) (Gupta et al. 2010; Mollenkamp and Kasten
28 2009). Governance processes and policy instruments supporting these characteristics are context
29 specific (*medium evidence, high agreement*) (Biermann 2007; Gunderson and Holling 2001; Hurlbert
30 and Gupta 2017; Bastos Lima et al. 2017a; Gupta et al. 2013a; Mollenkamp and Kasten 2009; Nelson
31 et al. 2010; Olsson et al. 2006; Ostrom 2011; Pahl-Wostl 2009; Verweij et al. 2006; Weick and
32 Sutcliffe 2001).

33 Consideration of these indicators is important when implementing climate change mitigation
34 instruments. For example, a ‘Variety,’ redundancy, or duplication of climate mitigation policy
35 instruments is an important consideration for meeting Paris Commitments. Given 58% of EU
36 emissions are outside of the EU Emissions Trading system, implementation of a ‘redundant’ carbon
37 tax may add co-benefits (Baranzini et al. 2017). Further, a carbon tax phased in over time through a
38 schedule of increases allows for ‘Learning.’ The tax revenues could be earmarked to finance
39 additional climate change mitigation and or redistributed to achieve the indicator of ‘Fair Governance
40 - Equity’. It is recommended that carbon pricing measures be implemented using information sharing
41 and communication devices to enable public acceptance, openness, provide measurement and
42 accountability (Baranzini et al. 2017; Siegmeier et al. 2018).

43 The impact of flood on a socio-ecological system is reduced with the governance indicator of both
44 leadership and resources (Emerson and Gerlak 2014). ‘Leadership’ pertains to a broad set of
45 stakeholders that facilitate adaptation (and might include scientists and leaders in NGOs) and those
46 that respond to flood in an open, inclusive, and fair manner identifying the most pressing issues and
47 actions needed. Resources are required to support this leadership and includes upfront financial
48 investment in human capital, technology, and infrastructure (Emerson and Gerlak 2014).

1 Policy instruments advancing the indicator of ‘Participation’ in community forest management
 2 include favourable loans, tax measures, and financial support to catalyse entrepreneurial leadership,
 3 and build in rewards for supportive and innovative elites to reduce elite capture and ensure more
 4 inclusive participation (Duguma et al. 2018) (see 7.6.4).

5 **Table 7.8 Institutional Dimensions or Indicators of Adaptive Governance**
 6 **This table represents a summation of characteristics, evaluative criteria, elements, indicators or**
 7 **institutional design principles that advance adaptive governance**

| Indicators/Inst itutional Dimensions | Description | References |
|--------------------------------------|---|--|
| Variety | Room for a variety of problem frames reflecting different opinions and problem definitions | (Biermann 2007; |
| | Participation. Involving different actors at different levels, sectors, and dimensions | Gunderson and Holling 2001; |
| | Availability of a wide range or diversity of policy options to address a particular problem | |
| | Redundancy or duplication of measures, back-up systems | Hurlbert and Gupta 2017; |
| Learning | Trust | Bastos Lima et al. 2017a; |
| | Single loop learning or ability to improve routines based on past experience | Gupta, J., van der Grijp, N., Kuik 2013; |
| | Double loop learning or changed underlying assumptions of institutional patterns | |
| | Discussion of doubts (openness to uncertainties, monitoring and evaluation of policy experiences) | |
| Room for autonomous change | Institutional memory (monitoring and evaluation of policy experiences over time) | |
| | Continuous access to information (data institutional memory and early warning systems) | Mollenkamp and Kasten 2009; |
| | Acting according to plan (especially in relation to disasters) | Nelson et al. 2010; |
| Leadership | Capacity to improvise (in relation to self-organisation and fostering social capital) | Olsson et al. 2006; |
| | Visionary (Long term and reformist) | |
| | Entrepreneurial which leads by example | Ostrom 2011; |
| Resources | Collaborative | |
| | Authority resources or legitimate forms of power | Pahl-Wostl 2009; |
| | Human resources of expertise, knowledge and labour | |
| Fair governance | Financial resources | Verweij et al. 2006; |
| | Legitimacy or public support | |
| | Equity in relation to institutional fair rules | Weick and Sutcliffe 2001) |
| | Responsiveness to society | |
| | Accountability in relation to procedures | |

8 **7.6.7. Inclusive Governance for Sustainable Development**

9 Many sustainable development efforts fail because of lack of attention to societal issues including
 10 inequality, discrimination, social exclusion and marginalisation (see Cross-Chapter Box 11: Gender in
 11 this chapter) (Arts 2017a). However, the human rights based approach of the 2030 Agenda and
 12 Sustainable Development Goals commits to leaving no one behind (Arts 2017b). Inclusive
 13 governance focuses attention in issues of equity and the human rights based approach for
 14 development as it includes social, ecological and relational components used for assessing access to,
 15 as well as the allocations of rights, responsibilities and risks with respect to social and ecological
 16 resources (medium agreement) (Gupta and Pouw 2017).

17 Governance processes that are inclusive of all people in decision making and management of land, are
 18 better able to make decisions addressing trade offs of sustainable development (Gupta et al. 2015) and
 19 achieve SDGs focusing on social and ecological inclusiveness (Gupta and Vegelin 2016). Citizen
 20 engagement is important in enhancing natural resource service delivery by citizen inclusion in
 21 management and governance decisions (see 7.5.5). In governing natural resources, focus is now not
 22 only on rights of citizens in relation to natural resources, but also on citizen obligations,
 23 responsibilities (Karar and Jacobs-Mata 2016; Chaney and Fevre 2001), feedback and learning
 24 processes (Tàbara et al. 2010). In this respect, citizen engagement is also an imperative particularly
 25 for analysing and addressing aggregated informal coping strategies of local residents in developing

1 countries, which are important drivers of natural resource depletions (but often overlooked in a
2 conventional policy development processes in natural resource management) (Ehara et al. 2018).

3 Inclusive adaptive governance makes important contributions to the management of risk. Inclusive
4 governance concerning risk integrates people's knowledge and values by involving them in decision
5 making processes where they are able to contribute their respective knowledge and values to make
6 effective, efficient, fair, and morally acceptable decisions (Renn and Schweizer 2009). Representation
7 in decision making would include major actors - government, economic sectors, the scientific
8 community and representatives of civil society (Renn and Schweizer 2009). Inclusive governance
9 focuses attention on the well being and meaningful participation in decision making of the poorest (in
10 income), vulnerable (in terms of age, gender, and location), and the most marginalised and is
11 inclusive of all knowledges (Gupta et al. 2015).

12 13 **7.7. Key uncertainties and knowledge gaps**

14 Uncertainties in land, society and climate change processes are outlined in 7.2 and Chapter 1. This
15 chapter has reviewed literature on risks arising from GHG Fluxes, climate change, land degradation,
16 desertification and food security, policy instruments responding to these risks, as well as decision
17 making and adaptive climate and land governance, in the face of uncertainty.

18 More research is required to understand the complex interconnections of land, climate, water, society,
19 ecosystem services and food, including:

- 20 • New models that allow incorporation of considerations of justice, inequality and human
21 agency in socio-environmental systems;
- 22 • Understanding how policy instruments and response options interact and augment or reduce
23 risks in relation to acute shocks and slow-onset climate events;
- 24 • Understanding how response options, policy, and instrument portfolios can reduce or augment
25 the cascading impacts of land, climate and food security and ecosystem service interactions
26 through different domains such as health, livelihoods, and infrastructure, especially in relation
27 to non-linear and tipping-point changes in natural and human systems.
- 28 • Consideration of trade-offs and synergies in climate, land, water, ecosystem services and food
29 policies;
- 30 • The impacts of increasing use of land due to climate mitigation measures such as BECCS,
31 carbon centric afforestation/REDD+ and their impacts on human conflict, livelihoods and
32 displacement;
- 33 • Understanding how different land tenure systems, both formal and informal, and the land
34 policies and administration systems that support them, can constrain or facilitate climate
35 adaptation and mitigation: and on how forms of climate action can enhance or undermine land
36 tenure security and land justice.
- 37 • Expanding understanding of barriers to implementation of land-based climate policies at all
38 levels from the local to the global, including methods for monitoring and documenting
39 corruption, misappropriation and elite capture in climate action;
- 40 • Identifying characteristics and attributes signalling impending socio-ecological tipping points
41 and collapse;
- 42 • Understanding the full cost of climate change in the context of disagreement on accounting
43 for climate change interactions and their impact on society, as well as issues of valuation, and
44 attribution uncertainties across generations;
- 45 • New models and Earth observation to understand complex interactions described in this
46 section.

- The impacts, monitoring, effectiveness, and appropriate selection of certification and standards for sustainability (see 7.4.6.3) (ISEAL Alliance; Stattman et al. 2018) and the effectiveness of its implementation through the landscape governance approach (Pacheco et al. 2016) (see 7.6.3).

Actions to mitigate climate change are rarely evaluated in relation to impact on adaptation, SDGs, and trade-offs with food security. For instance, there is a gap in knowledge in the optimal carbon pricing or emission trading scheme together with monitoring, reporting and verification system for agricultural emissions that will advance GHG reductions, food security, and sustainable land management. Better understanding is needed of the triggers and leveraging actions that build sustainable development and sustainable land management, as well as the effective organisation of the science and society interaction jointly shaping policies in the future. What societal interaction in the future will form inclusive and equitable governance processes and achieve inclusive just governance institutions including. Land tenure?

As there is a significant gap in NDCs and achieving commitments to keep global warming well below 2°C (7.4.4.1), governments might consider evaluating national, regional, and local gaps in knowledge surrounding response options, policy instruments portfolios, and sustainable land management supporting the achievement of NDCs in the face of land and climate change.

Frequently Asked Questions

FAQ 7.1 How can indigenous knowledge and local knowledge inform land-based mitigation and adaptation options?

Indigenous knowledge (IK) refers to the understandings, skills and philosophies developed by societies with long histories of interaction with their natural surroundings. Local knowledge (LK) refers to the understandings and skills developed by individuals and populations, specific to the place where they live. These forms of knowledge are often highly context-specific and embedded in local institutions, providing biological and ecosystem knowledge with landscape information. This means they can contribute to effective land management, predictions of natural disasters and identification of longer-term climate changes, for example, and IK can be particularly useful where formal data collection on environmental conditions may be sparse. IK and LK are often dynamic, with knowledge holders often experimenting with mixes of local and scientific approaches. Water management, soil fertility practices, grazing systems, restoration and sustainable harvesting of forests, and ecosystem based-adaptation are many of the land management practices often informed by IK and LK. LK can also be used as an entry point for climate adaptation by balancing past experiences with new ways to cope. To be effective, initiatives need to take into account the differences in power between the holders of different types of knowledge. For example, including indigenous and/or local people in programmes related to environmental conservation, formal education, land management planning and security tenure rights is key to facilitate climate change adaptation. Formal education is necessary to enhance adaptive capacity of IK and LK since some researchers have suggested these knowledge systems may become less relevant in certain areas where the rate of environmental change is rapid and the transmission of IK and LK between generations is becoming weaker.

FAQ 7.2 What are the main barriers to and opportunities for land-based responses to climate change?

Land-based responses to climate change can be mitigation (e.g., renewable energy, vegetation or crops for biofuels, afforestation) or adaptation (e.g., change in cropping pattern, less water intensive

1 crops in response to moisture stress), or adaptation with mitigation co-benefits (e.g., dietary shifts,
2 new uses for invasive tree-species, siting solar farms on highly degraded land). Productive land is an
3 increasingly scarce resource under climate change. In the absence of adequate deep mitigation in the
4 less land intensive energy sector, competition for land and water for mitigation and for other sectors
5 such as food security, ecosystem services and biodiversity conservation could become a source of
6 conflict and a barrier to land-based responses.

7 Barriers to land-based mitigation include opposition due to real and perceived trade-offs between land
8 for mitigation and food security and ecosystem services. These can arise due to absence of or
9 uncertain land and water rights. Significant upscaling of mitigation requires dedicated (normally
10 land-based) sources in addition to use of wastes and residues. This requires high land use intensity
11 compared to other mitigation options that in turn place greater demands on governance. A key
12 governance mechanism that has emerged in response to such concerns, especially during the past
13 decade are standards and certification systems that include food security and land and water rights in
14 addition to general criteria or indicators related to sustainable use of land and biomass with an
15 emphasis on participatory approaches. Other governance responses include linking land based
16 mitigation (e.g., forestry) to secure tenure and support for local livelihoods. A barrier to land-based
17 mitigation is our choice of development pathway. Our window of opportunity/ whether or not we face
18 barriers or opportunities to land based mitigation depends on socio-economic decisions or
19 pathways. If we have high population growth and resource intensive consumption (i.e., SSP3) we will
20 have more barriers. High population and low land use regulation results in less available space for
21 land based mitigation. But if we have the opposite trends (SSP1) we can have more opportunities.

22 Other barriers can arise when in the short term adaptation to a climate stress (eg increased dependence
23 on ground-water during droughts) can become unsustainable in the longer term and become a
24 maladaptation. Policies and approaches that lead to land management that synergises multiple
25 ecosystem services and reduce trade-offs could find greater acceptance and enjoy more success.

26 Opportunities to obtain benefits or synergies from land-based mitigation and adaptation arise
27 especially from their relation to the land availability and the demand for such measures in rural areas
28 that may otherwise lack incentives for investment in infrastructure, livelihoods and institutional
29 capacity. After decades of urbanisation around the world facilitated by significant investment in urban
30 infrastructure and centralised energy and agricultural systems, rural areas have been somewhat
31 neglected even as farmers in these areas provide critical food and materials needed for urban areas. As
32 land and biomass becomes more valuable, there will be benefits for farmers, forest owners and
33 associated service providers as they diversify away from feed and feed into economic activities
34 supporting bioenergy, value-added products, preservation of biodiversity and carbon sequestration
35 (storage).

36 A related opportunity for benefits is the potentially positive transformation in rural and peri-urban
37 landscapes that could be facilitated by investments that prioritise more effective management of
38 ecosystem services and conservation of water, energy, nutrients and other resources that have been
39 priced too low in relation to their environmental or ecological value. Multifunctional landscapes
40 supplying food, feed, fiber and fuel to both local and urban communities in combination with reduced
41 waste and healthier diets could restore the role of rural producers as stewards of resources rather than
42 providing food at the lowest possible price. Some of these landscape transformations will function as
43 both mitigation and adaptation responses by increasing resilience even as they provide value-added
44 bio-based products.

45 Governments can introduce a variety of regulations and economic instruments (taxes, incentives) to
46 encourage citizens, communities and societies to adopt sustainable land management practices with
47 further benefits in addition to mitigation. Windows of opportunity for redesigning and
48 implementing mitigation and adaptation can arise in the aftermath of a major disaster or extreme

1 climate event. They can also arise when collective action and citizen science motivate voluntary
2 shifts in lifestyles supported by supportive top-down policies.

5 References

6 Aalto, J., M. Kämäräinen, M. Shodmonov, N. Rajabov, and A. Venäläinen, 2017: Features of
7 Tajikistan's past and future climate. *Int. J. Climatol.*, **37**, 4949–4961, doi:10.1002/joc.5135.

8 Abatzoglou, J. T., and A. P. Williams, 2016: Impact of anthropogenic climate change on wildfire
9 across western US forests. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1607171113.

10 ———, ———, and R. Barbero, 2019a: Global Emergence of Anthropogenic Climate Change in Fire
11 Weather Indices. *Geophys. Res. Lett.*, **46**, 326–336, doi:10.1029/2018GL080959.
12 <http://doi.wiley.com/10.1029/2018GL080959> (Accessed April 14, 2019).

13 ———, ———, and ———, 2019b: Global Emergence of Anthropogenic Climate Change in Fire Weather
14 Indices. *Geophys. Res. Lett.*, **46**, 326–336, doi:10.1029/2018GL080959.

15 Abbasi, T., and S. A. Abbasi, 2011: Small hydro and the environmental implications of its extensive
16 utilization. *Renew. Sustain. energy Rev.*, **15**, 2134–2143.

17 Abbott, J., L. M. Campbell, C. J. Hay, T. F. Naesje, A. Ndumba, and J. Purvis, 2007: Rivers as
18 resources, rivers as borders: community and transboundary management of fisheries in the
19 Upper Zambezi River floodplains. *Can. Geogr. / Le Géographe Can.*, **51**, 280–302,
20 doi:10.1111/j.1541-0064.2007.00179.x. <http://doi.wiley.com/10.1111/j.1541-0064.2007.00179.x>
21 (Accessed November 4, 2018).

22 Abi-Samra, N. C., and W. P. Malcolm, 2011: Extreme weather effects on power systems. *IEEE Power
23 and Energy Society General Meeting*.

24 Abid, M., U. A. Schneider, and J. Scheffran, 2016: Adaptation to climate change and its impacts on
25 food productivity and crop income: Perspectives of farmers in rural Pakistan. *J. Rural Stud.*,
26 doi:10.1016/j.jrurstud.2016.08.005.

27 Achat, D. L., M. Fortin, G. Landmann, B. Ringeval, and L. Augusto, 2015a: Forest soil carbon is
28 threatened by intensive biomass harvesting. *Sci. Rep.*, **5**, 15991, doi:10.1038/srep15991.
29 <http://www.nature.com/articles/srep15991> (Accessed April 14, 2019).

30 ———, ———, ———, ———, and ———, 2015b: Forest soil carbon is threatened by intensive biomass
31 harvesting. *Sci. Rep.*, **5**, 15991, doi:10.1038/srep15991.

32 Adam, C., F. Taillandier, E. Delay, O. Plattard, and M. Toumi, 2016: SPRITE – Participatory
33 Simulation for Raising Awareness About Coastal Flood Risk on the Oleron Island. Springer,
34 Cham, 33–46 http://link.springer.com/10.1007/978-3-319-47093-1_4 (Accessed December 12,
35 2017).

36 Adano, W., and F. Daudi, 2012: Link between Climate change, Conflict and Governance in Africa.
37 *Inst. Secur. Stud.*, **234**.

38 Adano, W. R., T. Dietz, K. Witsenburg, and F. Zaal, 2012: Climate change, violent conflict and local
39 institutions in Kenya's drylands. *J. Peace Res.*, **49**, 65–80, doi:10.1177/0022343311427344.
40 <http://journals.sagepub.com/doi/10.1177/0022343311427344> (Accessed November 3, 2018).

41 Adaptation Sub-committee, 2013: Managing the land in a changing climate Chapter 5: Regulating
42 services - coastal habitats. 92–107.

43 Adaptation to Climate Change Team, 2013: *Summary for Decision Makers. Climate Change
44 Adaptation and Canada's Crop and Food Supply*. B.C. Canada, [http://act-adapt.org/wp-
45 content/uploads/2013/07/07-13-CFS-Summary-WEB.pdf](http://act-adapt.org/wp-content/uploads/2013/07/07-13-CFS-Summary-WEB.pdf) (Accessed October 3, 2018).

- 1 Adger, W. N., T. Quinn, I. Lorenzoni, C. Murphy, and J. Sweeney, 2013: Changing social contracts in
2 climate-change adaptation. *Nat. Clim. Chang.*, doi:10.1038/nclimate1751.
- 3 Admin, 2017: Follow India's lead: increase waterway freight, reduce emissions and accidents.
4 *Political Concern*, October [https://politicalcleanup.wordpress.com/2017/10/23/follow-indias-](https://politicalcleanup.wordpress.com/2017/10/23/follow-indias-lead-waterway-freight-reducing-emissions-and-accidents/)
5 [lead-waterway-freight-reducing-emissions-and-accidents/](https://politicalcleanup.wordpress.com/2017/10/23/follow-indias-lead-waterway-freight-reducing-emissions-and-accidents/).
- 6 Adnan, S., 2013: Land grabs and primitive accumulation in deltaic Bangladesh: Interactions between
7 neoliberal globalization, state interventions, power relations and peasant resistance. *J. Peasant*
8 *Stud.*, **40**, 87–128, doi:10.1080/03066150.2012.753058.
- 9 Adu, M., D. Yawson, F. Armah, E. Abano, and R. Quansah, 2018: Systematic review of the effects of
10 agricultural interventions on food security in northern Ghana. *PLoS One*, **13**,
11 <https://doi.org/10.1371/journal.pone.0203605>.
- 12 Aeschbach-Hertig, W., and T. Gleeson, 2012: Regional strategies for the accelerating global problem
13 of groundwater depletion. *Nat. Geosci.*, **5**, 853.
- 14 Agarwal, B., 2010: *Gender and green governance: The political economy of women's presence within*
15 *and beyond community forestry*. Oxford University Press., Oxford.,
16 [http://www.oxfordscholarship.com/view/10.1093/acprof:oso/9780199569687.001.0001/acprof-](http://www.oxfordscholarship.com/view/10.1093/acprof:oso/9780199569687.001.0001/acprof-9780199569687)
17 [9780199569687](http://www.oxfordscholarship.com/view/10.1093/acprof:oso/9780199569687.001.0001/acprof-9780199569687).
- 18 Agarwal, B., 2018a: Gender equality, food security and the sustainable development goals. *Curr.*
19 *Opin. Environ. Sustain.*, **34**, 26–32, doi:10.1016/j.cosust.2018.07.002.
20 <https://linkinghub.elsevier.com/retrieve/pii/S1877343517302415> (Accessed April 2, 2019).
- 21 —, 2018b: Gender equality, food security and the sustainable development goals. *Curr. Opin.*
22 *Environ. Sustain.*, **34**, 26–32, doi:10.1016/j.cosust.2018.07.002.
- 23 Aggarwal, A., 2011: Implementation of Forest Rights Act, changing forest landscape, and “politics of
24 REDD+” in India. *J. Resour. Energy Dev.*, **8**, 131–148.
- 25 Agrawal, A., A. Chhatre, and R. Hardin, 2008: Changing governance of the world's forests. *Science*
26 *(80-)*, **320**, 1460–1462.
- 27 Agriculture Technical Advisory Group, 2009: *Point of obligation designs and allocation*
28 *methodologies for agriculture and the New Zealand Emissions Trading Scheme*. Wellington,
29 <https://www.parliament.nz/resource/0000077853>.
- 30 Aguilar-Støen, M., 2017: Better Safe than Sorry? Indigenous Peoples, Carbon Cowboys and the
31 Governance of REDD in the Amazon. *Forum Dev. Stud.*, **44**, 91–108,
32 doi:10.1080/08039410.2016.1276098.
33 <https://www.tandfonline.com/doi/full/10.1080/08039410.2016.1276098> (Accessed October 15,
34 2018).
- 35 Aha, B., and J. Z. Ayitey, 2017: Biofuels and the hazards of land grabbing: Tenure (in)security and
36 indigenous farmers' investment decisions in Ghana. *Land use policy*, **60**, 48–59,
37 doi:10.1016/J.LANDUSEPOL.2016.10.012.
38 <https://www.sciencedirect.com/science/article/pii/S0264837715302180> (Accessed April 9,
39 2019).
- 40 Akhtar-Schuster, M., L. C. Stringer, A. Erlewein, G. Metternicht, S. Minelli, U. Safriel, and S.
41 Sommer, 2017: Unpacking the concept of land degradation neutrality and addressing its
42 operation through the Rio Conventions. *J. Environ. Manage.*,
43 doi:10.1016/j.jenvman.2016.09.044.
- 44 Alam, K., 2015: Farmers' adaptation to water scarcity in drought-prone environments: A case study of
45 Rajshahi District, Bangladesh. *Agric. Water Manag.*, doi:10.1016/j.agwat.2014.10.011.
- 46 Albers, R. A. W., and Coauthors, 2015: Overview of challenges and achievements in the climate
47 adaptation of cities and in the Climate Proof Cities program. *Building and Environment*.

- 1 Albert, C., J. Hauck, N. Buhr, and C. von Haaren, 2014: What ecosystem services information do
2 users want? Investigating interests and requirements among landscape and regional planners in
3 Germany. *Landsc. Ecol.*, **29**, 1301–1313, doi:10.1007/s10980-014-9990-5.
- 4 Albrecht, T. R., A. Crootof, and C. A. Scott, 2018: The water-energy-food nexus: A comprehensive
5 review of nexus-specific methods. *Environ. Res. Lett.*, doi:10.1088/1748-9326/aaa9c6.
- 6 Aldy, J. ., A. . Krupnick, R. . Newell, I. . Parry, and W. A. Pizer, 2010: Designing climate mitigation
7 policy. *J. Econ. Lit.*, **48**, 903–934.
- 8 Aldy, J. E., and R. N. Stavins, 2012: The Promise and Problems of Pricing Carbon. *J. Environ. Dev.*,
9 **21**, 152–180, doi:10.1177/1070496512442508.
10 <http://journals.sagepub.com/doi/10.1177/1070496512442508> (Accessed February 25, 2019).
- 11 Alexander, C., and Coauthors, 2011a: Linking Indigenous and Scientific Knowledge of Climate
12 Change. *Bioscience*, **61**, 477–484, doi:10.1525/bio.2011.61.6.10.
13 <https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2011.61.6.10> (Accessed
14 May 27, 2018).
- 15 ———, and Coauthors, 2011b: Linking Indigenous and Scientific Knowledge of Climate Change.
16 *Bioscience*, **61**, 477–484, doi:10.1525/bio.2011.61.6.10.
17 <https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2011.61.6.10> (Accessed
18 January 2, 2018).
- 19 Alexander, K. A., and Coauthors, 2015a: What Factors Might Have Led to the Emergence of Ebola in
20 West Africa? *PLoS Negl. Trop. Dis.*, **9**, doi:10.1371/journal.pntd.0003652.
- 21 Alexander, P., M. D. A. Rounsevell, C. Dislich, J. R. Dodson, K. Engström, and D. Moran, 2015b:
22 Drivers for global agricultural land use change: The nexus of diet, population, yield and
23 bioenergy. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2015.08.011.
- 24 Alimi, T. O., D. O. Fuller, W. A. Qualls, S. V. Herrera, M. Arevalo-Herrera, M. L. Quinones, M. V.
25 G. Lacerda, and J. C. Beier, 2015: Predicting potential ranges of primary malaria vectors and
26 malaria in northern South America based on projected changes in climate, land cover and human
27 population. *Parasit. Vectors*, **8**, 431, doi:10.1186/s13071-015-1033-9.
28 <http://www.parasitesandvectors.com/content/8/1/431> (Accessed December 29, 2017).
- 29 Alkire, S., R. Meinzen-Dick, A. Peterman, A. Quisumbing, G. Seymour, and A. Vaz, 2013a: The
30 Women’s Empowerment in Agriculture Index. *World Dev.*, **52**, 71–91,
31 doi:10.1016/j.worlddev.2013.06.007. <http://dx.doi.org/10.1016/j.worlddev.2013.06.007>
32 (Accessed April 2, 2019).
- 33 ———, ———, ———, ———, ———, and ———, 2013b: The Women’s Empowerment in Agriculture Index.
34 *World Dev.*, **52**, 71–91, doi:10.1016/j.worlddev.2013.06.007.
- 35 Allan, T., M. Keulertz, and E. Woertz, 2015: The water–food–energy nexus: an introduction to nexus
36 concepts and some conceptual and operational problems. *International Journal of Water*
37 *Resources Development*.
- 38 Allen, C. D., and Coauthors, 2010: A global overview of drought and heat-induced tree mortality
39 reveals emerging climate change risks for forests. *For. Ecol. Manage.*, **259**, 660–684,
40 doi:10.1016/j.foreco.2009.09.001. <https://hal.archives-ouvertes.fr/hal-00457602/document>
41 (Accessed May 17, 2018).
- 42 Allen, C. R., J. J. Fontaine, K. L. Pope, and A. S. Garmestani, 2011: Adaptive management for a
43 turbulent future. *J. Environ. Manage.*, **92**, 1339–1345, doi:10.1016/j.jenvman.2010.11.019.
- 44 Allen, M. R., J. S. Fuglestedt, K. P. Shine, A. Reisinger, R. T. Pierrehumbert, and P. M. Forster,
45 2016: New use of global warming potentials to compare cumulative and short-lived climate
46 pollutants. *Nat. Clim. Chang.*, **6**, 773. <http://dx.doi.org/10.1038/nclimate2998>.
- 47 Allison, E. H., and Coauthors, 2009: Vulnerability of national economies to the impacts of climate
48 change on fisheries. *Fish Fish.*, **10**, 173–196.

- 1 Allwood, J. M., V. Bosetti, N. K. Dubash, L. Gómez-Echeverri, and C. von Stechow, 2014: Glossary.
2 *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the*
3 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, O. Edenhofer et
4 al., Eds., Cambridge University Press, Cambridge and New York
5 https://www.ipcc.ch/pdf/assessment-report/ar5/wg3/ipcc_wg3_ar5_annex-i.pdf (Accessed
6 September 19, 2018).
- 7 Alrø, H. F., H. Møller, J. Læssøe, and E. Noe, 2016: Opportunities and challenges for multicriteria
8 assessment of food system sustainability. *Ecol. Soc.*, **21**, 38, doi:10.5751/ES-08394-210138.
9 <https://www.ecologyandsociety.org/vol21/iss1/art38/>.
- 10 Alston, J. M., and P. G. Pardey, 2014: Agriculture in the Global Economy. *J. Econ. Perspect.*,
11 doi:10.1257/jep.28.1.121.
- 12 Alston, M., 2006: The gendered impact of drought. *Rural Gender Relations: Issues And Case Studies*,
13 B.B. Bock and S. Shortall, Eds., Cambridge, Oxfordshire, UK, 165–180.
- 14 Alston, M., 2014: Gender mainstreaming and climate change. doi:10.1016/j.wsif.2013.01.016.
15 <http://dx.doi.org/10.1016/j.wsif.2013.01.016> (Accessed April 2, 2019).
- 16 ———, J. Clarke, K. Whittenbury, M. Alston, J. Clarke, and K. Whittenbury, 2018a: Contemporary
17 Feminist Analysis of Australian Farm Women in the Context of Climate Changes. *Soc. Sci.*, **7**,
18 16, doi:10.3390/socsci7020016. <http://www.mdpi.com/2076-0760/7/2/16> (Accessed April 2,
19 2019).
- 20 ———, ———, ———, ———, ———, and ———, 2018b: Contemporary Feminist Analysis of Australian Farm
21 Women in the Context of Climate Changes. *Soc. Sci.*, **7**, 16, doi:10.3390/socsci7020016.
- 22 Altieri, M. A., and C. I. Nicholls, 2017: The adaptation and mitigation potential of traditional
23 agriculture in a changing climate. *Clim. Change*, doi:10.1007/s10584-013-0909-y.
- 24 Altieri, M. A., C. I. Nicholls, A. Henao, and M. A. Lana, 2015: Agroecology and the design of
25 climate change-resilient farming systems. *Agron. Sustain. Dev.*, **35**, 869–890.
- 26 Alvarez, G., M. Elfving, and C. Andrade, 2016: REDD+ governance and indigenous peoples in Latin
27 America: the case of Suru Carbon Project in the Brazilian Amazon Forest. *Lat. Am. J. Manag.*
28 *Sustain. Dev.*, **3**, 133, doi:10.1504/LAJMSD.2016.083705.
29 <http://www.inderscience.com/link.php?id=83705> (Accessed March 9, 2019).
- 30 Alverson, K., and Z. Zommers, eds., 2018: *Resilience The Science of Adaptation to Climate Change*.
31 ELSEVIER SCIENCE BV, 360 pp.
- 32 Ammann, W. J., 2013: Disaster risk reduction. *Encyclopedia of Earth Sciences Series*.
- 33 van Andel, T., and L. G. Carneiro, 2013: Why urban citizens in developing countries use
34 traditional medicines: the case of Suriname. *Evid. Based. Complement. Alternat. Med.*, **2013**,
35 687197, doi:10.1155/2013/687197. <http://www.hindawi.com/journals/ecam/2013/687197/>
36 (Accessed April 9, 2019).
- 37 Andela, N., and Coauthors, 2017: A human-driven decline in global burned area. *Science*, **356**, 1356–
38 1362, doi:10.1126/science.aal4108. <http://www.ncbi.nlm.nih.gov/pubmed/28663495> (Accessed
39 April 14, 2019).
- 40 Anderegg, W. R. L., J. M. Kane, and L. D. L. Anderegg, 2013: Consequences of widespread tree
41 mortality triggered by drought and temperature stress. *Nat. Clim. Chang.*, **3**, 30–36,
42 doi:10.1038/nclimate1635. <http://www.nature.com/articles/nclimate1635> (Accessed May 23,
43 2018).
- 44 Anderson, C. M., C. B. Field, and K. J. Mach, 2017: Forest offsets partner climate-change mitigation
45 with conservation. *Front. Ecol. Environ.*, **15**, 359–365, doi:10.1002/fee.1515.
- 46 Anderson, E. P., M. C. Freeman, and C. M. Pringle, 2006: Ecological consequences of hydropower
47 development in Central America: impacts of small dams and water diversion on neotropical

- 1 stream fish assemblages. *River Res. Appl.*, **22**, 397–411.
- 2 Anderson, J. E., 2010: *Public policymaking: An Introduction*. Cengage Learning, Hampshire,.
- 3 Anderson, K., and G. Peters, 2016: The trouble with negative emissions. *Science (80-.)*, **354**, 182–
4 183, doi:10.1126/science.aah4567. <http://www.sciencemag.org/cgi/doi/10.1126/science.aah4567>
5 (Accessed December 8, 2017).
- 6 Anderson, S., J. Morton, and C. Toulmin, 2010: Climate Change for Agrarian Societies in Drylands:
7 Implications and future pathways. *Social Dimensions of Climate Change: Equity and*
8 *vulnerability in a warming world*, R. Mearns and A. Norton, Eds., The World Bank,
9 Washington, DC, 199–230.
- 10 Andersson, E., and S. Gabrielsson, 2012: “Because of poverty, we had to come together”: Collective
11 action for improved food security in rural Kenya and Uganda. *Int. J. Agric. Sustain.*, **10**, 245–
12 262, doi:10.1080/14735903.2012.666029.
- 13 Anenberg, S. C., L. W. Horowitz, D. Q. Tong, and J. J. West, 2010: An estimate of the global burden
14 of anthropogenic ozone and fine particulate matter on premature human mortality using
15 atmospheric modeling. *Environ. Health Perspect.*, doi:10.1289/ehp.0901220.
- 16 Angelsen, A., C. Martius, A. E. Duchelle, A. M. Larson, T. Thuy, and S. Wunder, 2018a: Conclusions
17 Lessons for the path to a transformational REDD+ Chapter 16. *Transforming REDD+: Lessons*
18 *and new directions*, A. Angelsen, C. Martius, V. De Sy, A. Duchelle, A. Larson, and T. Pham,
19 Eds., CIFOR, Bogor, Indonesia, 203–214 <https://www.cifor.org/library/7045> (Accessed March
20 9, 2019).
- 21 —, —, —, —, —, and —, 2018b: Conclusions Lessons for the path to a
22 transformational REDD+ Chapter 16. *Transforming REDD+: Lessons and new directions*, A.
23 Angelsen, C. Martius, V. De Sy, A. Duchelle, A. Larson, and T. Pham, Eds., CIFOR, Bogor,
24 Indonesia, 203–214.
- 25 Anguelovski, I., L. Shi, E. Chu, D. Gallagher, K. Goh, Z. Lamb, K. Reeve, and H. Teicher, 2016a:
26 Equity Impacts of Urban Land Use Planning for Climate Adaptation. *J. Plan. Educ. Res.*, **36**,
27 333–348, doi:10.1177/0739456X16645166.
28 <http://journals.sagepub.com/doi/10.1177/0739456X16645166> (Accessed February 25, 2019).
- 29 —, —, —, —, —, —, —, and —, 2016b: Equity Impacts of Urban Land Use
30 Planning for Climate Adaptation: Critical Perspectives from the Global North and South. *J.*
31 *Plan. Educ. Res.*, doi:10.1177/0739456X16645166.
- 32 Ansar, A., B. L. Caldecott, and J. Tilbury, 2013: Stranded assets and the fossil fuel divestment
33 campaign: what does divestment mean for the valuation of fossil fuel assets?
- 34 Anthoff, D., R. S. J. Tol, and G. Yohe, 2010: *Discounting for Climate Change*.
- 35 Antwi-Agyei, P., A. J. Dougill, and L. C. Stringer, 2015: Impacts of land tenure arrangements on the
36 adaptive capacity of marginalized groups: The case of Ghana’s Ejura Sekyedumase and Bongo
37 districts. *Land use policy*, **49**, 203–212, doi:10.1016/j.landusepol.2015.08.007.
- 38 AR5, 2014: *Glossary*. 1757-1776 pp.
- 39 Aragão, L. E. O. C., 2012: Environmental science: The rainforest’s water pump. *Nature*, **489**, 217.
- 40 Arakelyan, I., D. Moran, and A. Wreford, 2017: *Climate smart agriculture: A critical review*. F.
41 Nunan, Ed. Rout,.
- 42 Araujo Enciso, S. R., T. Fellmann, I. Pérez Dominguez, and F. Santini, 2016: Abolishing biofuel
43 policies: Possible impacts on agricultural price levels, price variability and global food security.
44 *Food Policy*, **61**, 9–26, doi:10.1016/J.FOODPOL.2016.01.007. [https://www.sciencedirect-](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0306919216000166)
45 [com.ezp.sub.su.se/science/article/pii/S0306919216000166](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0306919216000166) (Accessed May 23, 2018).
- 46 Archer, D., F. Almansi, M. DiGregorio, D. Roberts, D. Sharma, and D. Syam, 2014: Moving towards
47 inclusive urban adaptation: approaches to integrating community-based adaptation to climate

- 1 change at city and national scale. *Clim. Dev.*, doi:10.1080/17565529.2014.918868.
- 2 Argyris, C., 1999: *On organizational learning*. 1-4 pp.
- 3 Armeni, C., 2016: Participation in environmental decision-making: Reflecting on planning and
4 community benefits for major wind farms. *J. Environ. Law*, **28**, 415–441,
5 doi:10.1093/jel/eqw021.
- 6 Arnáez, J., N. Lana-Renault, T. Lasanta, P. Ruiz-Flaño, and J. Castroviejo, 2015: Effects of farming
7 terraces on hydrological and geomorphological processes. A review. *CATENA*, **128**, 122–134,
8 doi:10.1016/J.CATENA.2015.01.021.
9 <https://www.sciencedirect.com/science/article/pii/S0341816215000351> (Accessed April 8,
10 2019).
- 11 Arnell, N. W., and B. Lloyd-Hughes, 2014: The global-scale impacts of climate change on water
12 resources and flooding under new climate and socio-economic scenarios. *Clim. Change*, **122**,
13 127–140, doi:10.1007/s10584-013-0948-4. <http://link.springer.com/10.1007/s10584-013-0948-4>
14 (Accessed April 11, 2019).
- 15 Arora-Jonsson, S., 2011: Virtue and vulnerability: Discourses on women, gender and climate change.
16 *Glob. Environ. Chang.*, **21**, 744–751, doi:10.1016/j.gloenvcha.2011.01.005.
17 [http://www.montana.edu/empowering-women-in-ag/documents/articles-and-news/Virtue_and](http://www.montana.edu/empowering-women-in-ag/documents/articles-and-news/Virtue_and_vulnerability_Discourses_on_women_gender_and_climate_change.pdf)
18 [vulnerability_Discourses_on_women_gender_and_climate_change.pdf](http://www.montana.edu/empowering-women-in-ag/documents/articles-and-news/Virtue_and_vulnerability_Discourses_on_women_gender_and_climate_change.pdf) (Accessed April 2, 2019).
- 19 —, 2014: Forty years of gender research and environmental policy: Where do we stand? *Womens.*
20 *Stud. Int. Forum*, **47**, 295–308, doi:10.1016/J.WSIF.2014.02.009.
21 <https://www.sciencedirect.com/science/article/abs/pii/S0277539514000326> (Accessed April 2,
22 2019).
- 23 Arrow, K. J., and Coauthors, 2014: Should governments use a declining discount rate in project
24 analysis? *Rev. Environ. Econ. Policy*, doi:10.1093/reep/reu008.
- 25 Arts, K., 2017a: Inclusive sustainable development: a human rights perspective. *Curr. Opin. Environ.*
26 *Sustain.*, doi:10.1016/j.cosust.2017.02.001.
- 27 Arts, K., 2017b: Inclusive sustainable development: a human rights perspective. *Curr. Opin. Environ.*
28 *Sustain.*, doi:10.1016/j.cosust.2017.02.001.
- 29 Ashcroft, M., R. Austin, K. Barnes, D. MacDonald, S. Makin, S. Morgan, R. Taylor, and P. Scolley,
30 2016: Expert judgement. *Br. Actuar. J.*, **21**, 314–363, doi:10.1017/S1357321715000239.
- 31 Ashkenazy, A., T. Calvão Chebach, K. Knickel, S. Peter, B. Horowitz, and R. Offenbach, 2017:
32 Operationalising resilience in farms and rural regions – Findings from fourteen case studies. *J.*
33 *Rural Stud.*, doi:10.1016/J.JRURSTUD.2017.07.008. [https://www.sciencedirect-](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0743016716307161)
34 [com.ezp.sub.su.se/science/article/pii/S0743016716307161](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0743016716307161) (Accessed December 30, 2017).
- 35 Ashoori, N., D. A. Dzombak, and M. J. Small, 2015: Sustainability Review of Water-Supply Options
36 in the Los Angeles Region. *J. Water Resour. Plan. Manag.*, doi:10.1061/(ASCE)WR.1943-
37 5452.0000541.
- 38 Astrup, R., R. M. Bright, P. Y. Bernier, H. Genet, and D. A. Lutz, 2018: A sensible climate solution
39 for the boreal forest. *Nat. Clim. Chang.*,.
- 40 Atkinson, G., B. Groom, N. Hanley, and S. Mourato, 2018: Environmental Valuation and Benefit-
41 Cost Analysis in U.K. Policy. *J. Benefit-Cost Anal.*, **9**, 1–23, doi:10.1017/bca.2018.6.
- 42 Auclair, L., P. Baudot, D. Genin, B. Romagny, and R. Simenel, 2011: Patrimony for Resilience:
43 Evidence from the Forest Agdal in the Moroccan High Atlas Mountains. *Ecol. Soc.*, **16**, art24,
44 doi:10.5751/ES-04429-160424. <http://www.ecologyandsociety.org/vol16/iss4/art24/> (Accessed
45 April 8, 2019).
- 46 Awal, M. A., 2013: Social Safety Net , Disaster Risk Management and Climate Change Adaptation:
47 Examining Their Integration Potential in Bangladesh. *Int. J. Sociol. Study*, **1**, 62–72.

- 1 Ayamga, E. A., F. Kemausuor, and A. Addo, 2015a: Technical analysis of crop residue biomass
2 energy in an agricultural region of Ghana. *Resour. Conserv. Recycl.*, **96**, 51–60,
3 doi:10.1016/j.resconrec.2015.01.007. [https://www.sciencedirect-](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0921344915000087)
4 [com.ezp.sub.su.se/science/article/pii/S0921344915000087](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0921344915000087) (Accessed April 14, 2019).
- 5 ———, ———, and ———, 2015b: Technical analysis of crop residue biomass energy in an agricultural
6 region of Ghana. *Resour. Conserv. Recycl.*, **96**, 51–60, doi:10.1016/j.resconrec.2015.01.007.
- 7 Ayanlade, A., M. Radeny, and J. Morton, 2017: Comparing smallholder farmers’ perception of
8 climate change with meteorological data: A case study from southwestern Nigeria. *Weather*
9 *Clim. Extrem.*, **15**, 24–33.
- 10 Ayeb-Karlsson, S., K. van der Geest, I. Ahmed, S. Huq, and K. Warner, 2016: A people-centred
11 perspective on climate change, environmental stress, and livelihood resilience in Bangladesh.
12 *Sustain. Sci.*, doi:10.1007/s11625-016-0379-z.
- 13 Bacon, C. M., W. A. Sundstrom, M. E. Flores Gómez, V. Ernesto Méndez, R. Santos, B. Goldoftas,
14 and I. Dougherty, 2014: Explaining the “hungry farmer paradox”: Smallholders and fair trade
15 cooperatives navigate seasonality and change in Nicaragua’s corn and coffee markets. *Glob.*
16 *Environ. Chang.*, doi:10.1016/j.gloenvcha.2014.02.005.
- 17 Badola, R., S. C. Barthwal, and S. A. Hussain, 2013: Payment for ecosystem services for balancing
18 conservation and development in the rangelands of the Indian Himalayan region. *High-Altitude*
19 *Rangelands their Interfaces Hindu Kush Himalayas*, 175.
- 20 Bailey, R., 2013: The “Food Versus Fuel” Nexus. *The Handbook of Global Energy Policy*.
- 21 Bailis, R., R. Drigo, A. Ghilardi, and O. Masera, 2015: The carbon footprint of traditional woodfuels.
22 *Nat. Clim. Chang.*, **5**, 266–272, doi:10.1038/nclimate2491. [http://rembio.org.mx/wp-](http://rembio.org.mx/wp-content/uploads/2014/10/Bailis-et-al-2015-The-Carbon-Footprint-of-Traditional-Biofuels.pdf)
23 [content/uploads/2014/10/Bailis-et-al-2015-The-Carbon-Footprint-of-Traditional-Biofuels.pdf](http://rembio.org.mx/wp-content/uploads/2014/10/Bailis-et-al-2015-The-Carbon-Footprint-of-Traditional-Biofuels.pdf)
24 (Accessed May 18, 2018).
- 25 Bailis, R., Y. Wang, R. Drigo, A. Ghilardi, and O. Masera, 2017a: Getting the numbers right:
26 revisiting woodfuel sustainability in the developing world. *Environ. Res. Lett.*, **12**, 115002,
27 doi:10.1088/1748-9326/aa83ed. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=11/a=115002?key=crossref.55eef9f4eb061ce451b1c15af1f1f805)
28 [9326/12/i=11/a=115002?key=crossref.55eef9f4eb061ce451b1c15af1f1f805](http://stacks.iop.org/1748-9326/12/i=11/a=115002?key=crossref.55eef9f4eb061ce451b1c15af1f1f805) (Accessed April 2,
29 2019).
- 30 ———, ———, ———, ———, and ———, 2017b: Getting the numbers right: revisiting woodfuel
31 sustainability in the developing world. *Environ. Res. Lett.*, **12**, 115002, doi:10.1088/1748-
32 [9326/aa83ed](http://stacks.iop.org/1748-9326/12/i=11/a=115002?key=crossref.55eef9f4eb061ce451b1c15af1f1f805).
- 33 Baird, I. G., and P. Dearden, 2003: Biodiversity conservation and resource tenure regimes: a case
34 study from northeast Cambodia. *Environ. Manage.*, **32**, 541–550.
- 35 Baker, Dean; Jayadev, Arjun; Stiglitz, J., 2017: *Innovation, Intellectual Property, and Development:*
36 *A Better Set of Approaches for the 21st Century*. [http://ip-unit.org/wp-](http://ip-unit.org/wp-content/uploads/2017/07/IP-for-21st-Century-EN.pdf)
37 [content/uploads/2017/07/IP-for-21st-Century-EN.pdf](http://ip-unit.org/wp-content/uploads/2017/07/IP-for-21st-Century-EN.pdf) (Accessed October 15, 2018).
- 38 Di Baldassarre, G., M. Kooy, J. S. Kemerink, and L. Brandimarte, 2013: Towards understanding the
39 dynamic behaviour of floodplains as human-water systems. *Hydrol. Earth Syst. Sci.*, **17**, 3235–
40 3244, doi:10.5194/hess-17-3235-2013. <http://www.hydrol-earth-syst-sci.net/17/3235/2013/>
41 (Accessed May 18, 2018).
- 42 Balehegn, M., 2015: Unintended consequences: The ecological repercussions of land grabbing in sub-
43 Saharan Africa. *Environment*, **57**, 4–21, doi:10.1080/00139157.2015.1001687.
- 44 Ballard, H. L., C. G. H. Dixon, and E. M. Harris, 2017: Youth-focused citizen science: Examining the
45 role of environmental science learning and agency for conservation. *Biol. Conserv.*, **208**, 65–75,
46 doi:10.1016/j.biocon.2016.05.024.
- 47 Balooni, K., J. M. Pulhin, and M. Inoue, 2008: The effectiveness of decentralisation reforms in the
48 Philippines’s forestry sector. *Geoforum*, **39**, 2122–2131, doi:10.1016/j.geoforum.2008.07.003.

- 1 Bamberg, S., J. Rees, and S. Seebauer, 2015: Collective climate action: Determinants of participation
2 intention in community-based pro-environmental initiatives. *J. Environ. Psychol.*, **43**, 155–165,
3 doi:10.1016/J.JENVP.2015.06.006.
4 <https://www.sciencedirect.com/science/article/pii/S0272494415300190> (Accessed May 17,
5 2018).
- 6 —, J. H. Rees, and M. Schulte, 2018: 8 - Environmental protection through societal change: What
7 psychology knows about collective climate action—and what it needs to find out. *Psychology
8 and Climate Change*.
- 9 Banerjee, A., and Coauthors, 2015: A multifaceted program causes lasting progress for the very poor:
10 Evidence from six countries. *Science (80-.)*, **348**, doi:10.1126/science.1260799.
- 11 Baral, H., R. J. Keenan, S. K. Sharma, N. E. Stork, and S. Kasel, 2014: Economic evaluation of
12 ecosystem goods and services under different landscape management scenarios. *Land use policy*,
13 doi:10.1016/j.landusepol.2014.03.008.
- 14 Baranzini, A., J. C. J. M. van den Bergh, S. Carattini, R. B. Howarth, E. Padilla, and J. Roca, 2017:
15 Carbon pricing in climate policy: seven reasons, complementary instruments, and political
16 economy considerations. *Wiley Interdiscip. Rev. Clim. Chang.*, **8**, doi:10.1002/wcc.462.
- 17 Barbedo, J., M. Miguez, D. van der Horst, P. Carneiro, P. Amis, and A. Ioris, 2015: Policy
18 dimensions of land-use change in peri-urban floodplains: The case of paraty. *Ecol. Soc.*, **20**,
19 doi:10.5751/ES-07126-200105.
- 20 Barbier, E. B., 2011: *Pricing Nature*.
- 21 —, and A. T. Tesfaw, 2012: Can REDD+ Save the Forest? The Role of Payments and Tenure.
22 *Forests*, **3**, 881–895, doi:10.3390/f3040881. <http://www.mdpi.com/1999-4907/3/4/881>
23 (Accessed May 22, 2018).
- 24 Barkat, A., S. uz Zaman, S. Raihan, M. Rahman Chowdhury, and E. Director, 2001: *POLITICAL
25 ECONOMY OF KHAS LAND IN BANGLADESH*. [http://www.hdrc-
26 bd.com/admin_panel/images/notice/1389588575.3.cover.11.text.pdf](http://www.hdrc-bd.com/admin_panel/images/notice/1389588575.3.cover.11.text.pdf) (Accessed November 3,
27 2018).
- 28 Barnett, J., and S. O’Neill, 2010: Maladaptation. *Global Environmental Change*.
- 29 —, and J. P. Palutikof, 2014: The limits to adaptation: A comparative analysis. *Applied Studies in
30 Climate Adaptation*.
- 31 Barnett, J., S. Graham, C. Mortreux, R. Fincher, E. Waters, and A. Hurlimann, 2014: A local coastal
32 adaptation pathway. *Nat. Clim. Chang.*, doi:10.1038/nclimate2383.
- 33 Barnett, T. P., J. C. Adam, and D. P. Lettenmaier, 2005: Potential impacts of a warming climate on
34 water availability in snow-dominated regions. *Nature*, **438**, 303–309, doi:10.1038/nature04141.
35 <http://www.nature.com/articles/nature04141> (Accessed May 24, 2018).
- 36 Barrett, C. B., 2005: Rural poverty dynamics: development policy implications. *Agric. Econ.*, **32**, 45–
37 60, doi:10.1111/j.0169-5150.2004.00013.x.
- 38 Barrientos, A., 2011: Social protection and poverty. *Int. J. Soc. Welf.*, **20**, 240–249,
39 doi:10.1111/j.1468-2397.2011.00783.x.
- 40 Barros, F. S. M., and N. A. Honório, 2015: Deforestation and Malaria on the Amazon Frontier: Larval
41 Clustering of *Anopheles darlingi* (Diptera: Culicidae) Determines Focal Distribution of Malaria.
42 *Am. J. Trop. Med. Hyg.*, **93**, 939–953, doi:10.4269/ajtmh.15-0042.
43 <http://www.ajtmh.org/content/journals/10.4269/ajtmh.15-0042> (Accessed December 29, 2017).
- 44 Bastos Lima, M. G., and Coauthors, 2017a: The Sustainable Development Goals and REDD+:
45 assessing institutional interactions and the pursuit of synergies. *Int. Environ. Agreements Polit.
46 Law Econ.*, **17**, 589–606, doi:10.1007/s10784-017-9366-9.
- 47 Bastos Lima, M. G., I. J. Visseren-Hamakers, J. Braña-Varela, and A. Gupta, 2017b: A reality check

- 1 on the landscape approach to REDD+: Lessons from Latin America. *For. Policy Econ.*, **78**, 10–
2 20, doi:10.1016/J.FORPOL.2016.12.013.
3 <https://www.sciencedirect.com/science/article/pii/S1389934116301265> (Accessed October 15,
4 2018).
- 5 Bates, S., and P. Saint-Pierre, 2018: Adaptive Policy Framework through the Lens of the Viability
6 Theory: A Theoretical Contribution to Sustainability in the Anthropocene Era. *Ecol. Econ.*,
7 doi:10.1016/j.ecolecon.2017.09.007.
- 8 Batie, S. S., 2008: Wicked problems and applied economics. *Am. J. Agric. Econ.*, doi:10.1111/j.1467-
9 8276.2008.01202.x.
- 10 Baudron, F., M. Jaleta, O. Okitoi, and A. Tegegn, 2014: Conservation agriculture in African mixed
11 crop-livestock systems: Expanding the niche. *Agric. Ecosyst. Environ.*,
12 doi:10.1016/j.agee.2013.08.020.
- 13 Baulch, B., J. Wood, and A. Weber, 2006: Developing a social protection index for asia. *Dev. Policy*
14 *Rev.*, **24**, 5–29, doi:10.1111/j.1467-7679.2006.00311.x.
- 15 Baumgart-Getz, A., L. S. Prokopy, and K. Floress, 2012: Why farmers adopt best management
16 practice in the United States: A meta-analysis of the adoption literature. *J. Environ. Manage.*,
17 **96**, 17–25, doi:10.1016/j.jenvman.2011.10.006.
- 18 Bausch, J. ., L. Bojo'rquez-Tapia, and H. Eakin, 2014: Agro-environmental sustainability assessment
19 using multi-criteria decision analysis and system analysis. *Sustain. Sci.*, **9**, 303–319.
- 20 Bayisenge, J., 2018: From male to joint land ownership: Women's experiences of the land tenure
21 reform programme in Rwanda. *J. Agrar. Chang.*, **18**, 588–605, doi:10.1111/joac.12257.
22 <http://doi.wiley.com/10.1111/joac.12257> (Accessed April 2, 2019).
- 23 Bebbington, A. J., and Coauthors, 2018: Resource extraction and infrastructure threaten forest cover
24 and community rights. *Proc. Natl. Acad. Sci.*, **115**, 13164–13173,
25 doi:10.1073/PNAS.1812505115. <https://www.pnas.org/content/115/52/13164.short> (Accessed
26 April 9, 2019).
- 27 Beca Ltd, 2018: *Assessment of the administration costs and barriers of scenarios to mitigate*
28 *biological emissions from agriculture*.
- 29 Behera, B., B. Rahut, A. Jeetendra, and A. Ali, 2015a: Household collection and use of biomass
30 energy sources in South Asia. *Energy*, **85**, 468–480, doi:10.1016/j.energy.2015.03.059.
31 <http://dx.doi.org/10.1016/j.energy.2015.03.059> (Accessed April 14, 2019).
- 32 —, —, —, and —, 2015b: Household collection and use of biomass energy sources in South
33 Asia. *Energy*, **85**, 468–480, doi:10.1016/j.energy.2015.03.059.
- 34 Behnke, R., 1994: Natural Resource Management in Pastoral Africa. *Dev. Policy Rev.*, **12**, 5–28,
35 doi:10.1111/j.1467-7679.1994.tb00053.x. [http://doi.wiley.com/10.1111/j.1467-
36 7679.1994.tb00053.x](http://doi.wiley.com/10.1111/j.1467-7679.1994.tb00053.x) (Accessed May 22, 2018).
- 37 —, and C. Kerven, 2013: Counting the costs: replacing pastoralism with irrigated agriculture in the
38 Awash valley, north-eastern Ethiopia. *Pastoralism and development in Africa: dynamic changes*
39 *at the margins*, J. and S.I. Catley, A., Lind, Ed., Routledge, London, p. 49
40 <https://www.jstor.org/stable/resrep01236> (Accessed April 13, 2019).
- 41 Behrman, K. D., T. E. Juenger, J. R. Kiniry, and T. H. Keitt, 2015: Spatial land use trade-offs for
42 maintenance of biodiversity, biofuel, and agriculture. *Landsc. Ecol.*, doi:10.1007/s10980-015-
43 0225-1.
- 44 Belcher, B., M. Ruíz-Pérez, and R. Achdiawan, 2005: Global patterns and trends in the use and
45 management of commercial NTFPs: Implications for livelihoods and conservation. *World Dev.*,
46 **33**, 1435–1452, doi:10.1016/j.worlddev.2004.10.007.
47 <http://citeseerx.ist.psu.edu/viewdoc/download?doi=10.1.1.544.6523&rep=rep1&type=pdf>
48 (Accessed May 17, 2018).

- 1 Bellemare, M. F., 2015: Rising food prices, food price volatility, and social unrest. *Am. J. Agric.*
2 *Econ.*, doi:10.1093/ajae/aau038.
- 3 Below, T. B., K. D. Mutabazi, D. Kirschke, C. Franke, S. Sieber, R. Siebert, and K. Tscherning, 2012:
4 Can farmers' adaptation to climate change be explained by socio-economic household-level
5 variables? *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2011.11.012.
- 6 Ben-Ari, T., and D. Makowski, 2016: Analysis of the trade-off between high crop yield and low yield
7 instability at the global scale. *Environ. Res. Lett.*, **11**, doi:10.1088/1748-9326/11/10/104005.
- 8 Béné, C., B. Hersoug, and E. H. Allison, 2010: Not by rent alone: analysing the pro-poor functions of
9 small-scale fisheries in developing countries. *Dev. Policy Rev.*, **28**, 325–358.
- 10 —, and Coauthors, 2011: Testing resilience thinking in a poverty context: Experience from the
11 Niger River basin. *Glob. Environ. Chang.*, **21**, 1173–1184,
12 doi:10.1016/j.gloenvcha.2011.07.002.
13 <http://linkinghub.elsevier.com/retrieve/pii/S0959378011001075> (Accessed November 4, 2018).
- 14 —, S. Devereux, and R. Sabates-Wheeler, 2012: *Shocks and social protection in the Horn of*
15 *Africa: analysis from the Productive Safety Net programme in Ethiopia*. 1-120 pp.
- 16 Bennett, J. E., 2013: Institutions and governance of communal rangelands in South Africa. *African J.*
17 *Range Forage Sci.*, **30**, 77–83, doi:10.2989/10220119.2013.776634.
- 18 Bennett, N. J., and P. Dearden, 2014: From measuring outcomes to providing inputs: Governance,
19 management, and local development for more effective marine protected areas. *Mar. Policy*,
20 doi:10.1016/j.marpol.2014.05.005.
- 21 Bennich, T., S. Belyazid, T. Bennich, and S. Belyazid, 2017a: The Route to Sustainability—Prospects
22 and Challenges of the Bio-Based Economy. *Sustainability*, **9**, 887, doi:10.3390/su9060887.
23 <http://www.mdpi.com/2071-1050/9/6/887> (Accessed April 8, 2019).
- 24 —, —, —, and —, 2017b: The Route to Sustainability—Prospects and Challenges of the
25 Bio-Based Economy. *Sustainability*, **9**, 887, doi:10.3390/su9060887.
- 26 Bentz, B. J., and Coauthors, 2010: Climate Change and Bark Beetles of the Western United States and
27 Canada: Direct and Indirect Effects. *Bioscience*, **60**, 602–613, doi:10.1525/bio.2010.60.8.6.
28 <https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2010.60.8.6> (Accessed
29 April 14, 2019).
- 30 Benveniste, H., O. Boucher, C. Guivarch, H. Le Treut, and P. Criqui, 2018: Impacts of nationally
31 determined contributions on 2030 global greenhouse gas emissions: uncertainty analysis and
32 distribution of emissions. *Environ. Res. Lett.*, **13**, 014022, doi:10.1088/1748-9326/aaa0b9.
33 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=1/a=014022?key=crossref.0425f40986caf3ddddd8a8239f781d78b)
34 [9326/13/i=1/a=014022?key=crossref.0425f40986caf3ddddd8a8239f781d78b](http://stacks.iop.org/1748-9326/13/i=1/a=014022?key=crossref.0425f40986caf3ddddd8a8239f781d78b) (Accessed March
35 17, 2019).
- 36 van den Bergh, J. C. J. M., and W. J. W. Botzen, 2014: A lower bound to the social cost of CO₂
37 emissions. *Nat. Clim. Chang.*, doi:10.1038/nclimate2135.
- 38 Berhane, G., 2014: Can Social Protection Work in Africa? The Impact of Ethiopia's Productive Safety
39 Net Programme. *Econ. Dev. Cult. Change*, doi:10.1086/677753.
- 40 Berke, P. R., and M. R. Stevens, 2016: Land Use Planning for Climate Adaptation. *J. Plan. Educ.*
41 *Res.*, **36**, 283–289, doi:10.1177/0739456X16660714.
42 <http://journals.sagepub.com/doi/10.1177/0739456X16660714> (Accessed February 25, 2019).
- 43 Berkes, F., 2001: *Managing small-scale fisheries: alternative directions and methods*. IDRC,.
- 44 Berndes, G., S. Ahlgren, P. Börjesson, and A. L. Cowie, 2015: Bioenergy and Land Use Change-
45 State of the Art. *Advances in Bioenergy: The Sustainability Challenge*.
- 46 Bertram, M. Y., and Coauthors, 2018: Investing in non-communicable diseases: an estimation of the
47 return on investment for prevention and treatment services. *The Lancet*.

- 1 Bettini, G., and G. Gioli, 2016: Waltz with development: insights on the developmentalization of
2 climate-induced migration. *Migr. Dev.*, doi:10.1080/21632324.2015.1096143.
- 3 Bettini, Y., R. R. Brown, and F. J. de Haan, 2015: Exploring institutional adaptive capacity in
4 practice: Examining water governance adaptation in Australia. *Ecol. Soc.*, doi:10.5751/ES-
5 07291-200147.
- 6 Beuchelt, T. D., and L. Badstue, 2013: Gender, nutrition- and climate-smart food production:
7 Opportunities and trade-offs. *Food Secur.*, **5**, 709–721, doi:10.1007/s12571-013-0290-8.
- 8 Beuchle, R., R. C. Grecchi, Y. E. Shimabukuro, R. Seliger, H. D. Eva, E. Sano, and F. Achard, 2015:
9 Land cover changes in the Brazilian Cerrado and Caatinga biomes from 1990 to 2010 based on a
10 systematic remote sensing sampling approach. *Appl. Geogr.*, **58**, 116–127,
11 doi:10.1016/J.APGEOG.2015.01.017.
12 <https://www.sciencedirect.com/science/article/pii/S0143622815000284> (Accessed April 23,
13 2019).
- 14 Bevir, M., 2011: *The SAGE handbook of governance*.
- 15 Biagini, B., and A. Miller, 2013: Engaging the private sector in adaptation to climate change in
16 developing countries: Importance, status, and challenges. *Clim. Dev.*, **5**, 242–252,
17 doi:10.1080/17565529.2013.821053.
- 18 —, L. Kuhl, K. S. Gallagher, and C. Ortiz, 2014: Technology transfer for adaptation. *Nat. Clim.*
19 *Chang.*, **4**, 828–834, doi:10.1038/NCLIMATE2305. [https://www-nature-](https://www-nature-com.ezp.sub.su.se/articles/nclimate2305.pdf)
20 [com.ezp.sub.su.se/articles/nclimate2305.pdf](https://www-nature-com.ezp.sub.su.se/articles/nclimate2305.pdf) (Accessed May 22, 2018).
- 21 Biazin, B., G. Sterk, M. Temesgen, A. Abdulkedir, and L. Stroosnijder, 2012: Rainwater harvesting
22 and management in rainfed agricultural systems in sub-Saharan Africa – A review. *Phys. Chem.*
23 *Earth, Parts A/B/C*, **47–48**, 139–151, doi:10.1016/J.PCE.2011.08.015.
24 <https://www.sciencedirect.com/science/article/pii/S147470651100235X> (Accessed April 8,
25 2019).
- 26 Bierbaum, R., and A. Cowie, 2018: *Integration: to solve complex environmental problems. Scientific*
27 *and Technical Advisory Panel to the Global Environment Facility*. Washington D C,
28 <http://www.stapgef.org> (Accessed February 28, 2019).
- 29 Biermann, F., 2007: “Earth system governance” as a crosscutting theme of global change research.
30 *Glob. Environ. Chang.*, **17**, 326–337, doi:10.1016/j.gloenvcha.2006.11.010.
- 31 Biermann, F., and Coauthors, 2012: Science and government. Navigating the anthropocene:
32 improving Earth system governance. *Science*, **335**, 1306–1307, doi:10.1126/science.1217255.
33 <http://www.ncbi.nlm.nih.gov/pubmed/22422966> (Accessed December 21, 2017).
- 34 Biesbroek, G. R., R. J. Swart, T. R. Carter, C. Cowan, T. Henrichs, H. Mela, M. D. Morecroft, and D.
35 Rey, 2010: Europe adapts to climate change: Comparing National Adaptation Strategies. *Glob.*
36 *Environ. Chang.*, **20**, 440–450, doi:10.1016/j.gloenvcha.2010.03.005.
- 37 Biesbroek, R., B. G. Peters, and J. Tosun, 2018: Public Bureaucracy and Climate Change Adaptation.
38 *Rev. Policy Res.*, doi:10.1111/ropr.12316.
- 39 Biggs, E. M., E. L. Tompkins, J. Allen, C. Moon, and R. Allen, 2013: Agricultural adaptation to
40 climate change: observations from the Mid-Hills of Nepal. *Clim. Dev.*, **5**, 165–173,
41 doi:10.1080/17565529.2013.789791.
42 <http://www.tandfonline.com/doi/abs/10.1080/17565529.2013.789791> (Accessed April 23,
43 2019).
- 44 Biggs, E. M., and Coauthors, 2015: Sustainable development and the water-energy-food nexus: A
45 perspective on livelihoods. *Environ. Sci. Policy*, **54**, 389–397, doi:10.1016/j.envsci.2015.08.002.
46 <http://dx.doi.org/10.1016/j.envsci.2015.08.002>.
- 47 Biggs, H. C., J. K. Clifford-Holmes, S. Freitag, F. J. Venter, and J. Venter, 2017: Cross-scale
48 governance and ecosystem service delivery: A case narrative from the Olifants River in north-

- 1 eastern South Africa. *Ecosyst. Serv.*, doi:10.1016/j.ecoser.2017.03.008.
- 2 Binswanger, H. P., K. Deininger, and F. Gershon, 1995: Power, distortions, revolt, and reform in
3 agricultural land relations. *Handb. Dev. Econ.*, **3**, 2659–2772.
4 [http://documents.worldbank.org/curated/en/304261468764712147/Power-distortions-revolt-and-](http://documents.worldbank.org/curated/en/304261468764712147/Power-distortions-revolt-and-reform-in-agricultural-land-relations)
5 [reform-in-agricultural-land-relations](http://documents.worldbank.org/curated/en/304261468764712147/Power-distortions-revolt-and-reform-in-agricultural-land-relations) (Accessed November 3, 2018).
- 6 Birch, J. C., A. C. Newton, C. A. Aquino, E. Cantarello, C. Echeverría, T. Kitzberger, I.
7 Schiappacasse, and N. T. Garavito, 2010: Cost-effectiveness of dryland forest restoration
8 evaluated by spatial analysis of ecosystem services. *Proc. Natl. Acad. Sci.*, **107**, 21925–21930.
- 9 Bird, T. J., and Coauthors, 2014: Statistical solutions for error and bias in global citizen science
10 datasets. *Biol. Conserv.*, **173**, 144–154, doi:10.1016/J.BIOCON.2013.07.037.
11 <https://www.sciencedirect.com/science/article/pii/S0006320713002693> (Accessed October 2,
12 2018).
- 13 Bishop, I. D., C. J. Pettit, F. Sheth, and S. Sharma, 2013: Evaluation of Data Visualisation Options for
14 Land-Use Policy and Decision Making in Response to Climate Change. *Environ. Plan. B Plan.*
15 *Des.*, **40**, 213–233, doi:10.1068/b38159. <http://journals.sagepub.com/doi/10.1068/b38159>
16 (Accessed December 19, 2017).
- 17 Bizikova, L., D. Roy, D. Swanson, and M. Venema, Henry David, McCandless, 2013: The Water-
18 energy-food Security Nexus: Towards a Practical Planning and Decision-support Framework for
19 Landscape Investment and Risk Management. *Int. Inst. Sustain. Dev.*,.
- 20 Black, R., N. W. Arnell, W. N. Adger, D. Thomas, and A. Geddes, 2013: Migration, immobility and
21 displacement outcomes following extreme events. *Environ. Sci. Policy*,
22 doi:10.1016/j.envsci.2012.09.001.
- 23 Blackman, A., L. Corral, E. S. Lima, and G. P. Asner, 2017: Titling indigenous communities protects
24 forests in the Peruvian Amazon. *Proc. Natl. Acad. Sci.*, **114**, 4123–4128,
25 doi:10.1073/pnas.1603290114.
- 26 Le Blanc, D., 2015: Towards integration at last? The sustainable development goals as a network of
27 targets. *Sustain. Dev.*, **23**, 176–187.
- 28 Blanc, D. Le, C. Freire, and M. Vierro, 2017: Mapping the linkages between oceans and other
29 Sustainable Development Goals : A preliminary exploration. *DESA Work. Pap.*,.
- 30 Blanco-Canqui, H., and R. Lal, 2009a: Crop Residue Removal Impacts on Soil Productivity and
31 Environmental Quality. *CRC. Crit. Rev. Plant Sci.*, **28**, 139–163,
32 doi:10.1080/07352680902776507.
33 <http://www.tandfonline.com/doi/abs/10.1080/07352680902776507> (Accessed April 14, 2019).
- 34 ———, and ———, 2009b: Crop Residue Removal Impacts on Soil Productivity and Environmental
35 Quality. *CRC. Crit. Rev. Plant Sci.*, **28**, 139–163, doi:10.1080/07352680902776507.
- 36 Bloemen, P., M. Van Der Steen, and Z. Van Der Wal, 2018: Designing a century ahead: climate
37 change adaptation in the Dutch Delta. *Policy and Society*.
- 38 Blue, G., and J. Medlock, 2014: Public Engagement with Climate Change as Scientific Citizenship: A
39 Case Study of World Wide Views on Global Warming. *Sci. Cult. (Lond.)*, **23**, 560–579,
40 doi:10.1080/09505431.2014.917620.
- 41 Bodnár, F., B. de Steenhuijsen PETERS, and J. Kranen, 2011: Improving Food Security: A systematic
42 review of the impact of interventions in agricultural production, value chains, market regulation
43 and land security. *Minist. Foreign Aff. Netherlands*,.
- 44 Bodnar, P., C. Ott, R. Edwards, S. Hoch, E. F. McGlynn, and G. Wagner, 2018: Underwriting 1.5°C:
45 competitive approaches to financing accelerated climate change mitigation. *Clim. Policy*, **18**,
46 368–382, doi:10.1080/14693062.2017.1389687.
- 47 Bogaerts, M., L. Cirhigiri, I. Robinson, M. Rodkin, R. Hajjar, C. Costa Junior, and P. Newton, 2017:

- 1 Climate change mitigation through intensified pasture management: Estimating greenhouse gas
2 emissions on cattle farms in the Brazilian Amazon. *J. Clean. Prod.*, **162**, 1539–1550,
3 doi:10.1016/J.JCLEPRO.2017.06.130.
4 <https://www.sciencedirect.com/science/article/pii/S0959652617313008> (Accessed April 23,
5 2019).
- 6 Bogale, A., 2015a: Weather-indexed insurance: an elusive or achievable adaptation strategy to climate
7 variability and change for smallholder farmers in Ethiopia. *Clim. Dev.*, **7**, 246–256,
8 doi:10.1080/17565529.2014.934769.
9 <http://www.tandfonline.com/doi/full/10.1080/17565529.2014.934769> (Accessed May 16, 2018).
- 10 —, 2015b: Weather-indexed insurance: an elusive or achievable adaptation strategy to climate
11 variability and change for smallholder farmers in Ethiopia. *Clim. Dev.*, **7**, 246–256,
12 doi:10.1080/17565529.2014.934769.
- 13 Bogardi, J. J., D. Dudgeon, R. Lawford, E. Flinkerbusch, A. Meyn, C. Pahl-Wostl, K. Vielhauer, and
14 C. Vörösmarty, 2012: Water security for a planet under pressure: interconnected challenges of a
15 changing world call for sustainable solutions. *Curr. Opin. Environ. Sustain.*, **4**, 35–43.
- 16 Bohra-Mishra, P., and D. S. Massey, 2011: Environmental degradation and out-migration: New
17 evidence from Nepal. *Migration and climate change*.
- 18 Böhringer, C., J. C. Carbone, and T. F. Rutherford, 2012: Unilateral climate policy design: Ef fi
19 ciency and equity implications of alternative instruments to reduce carbon leakage. *Energy*
20 *Econ.*, **34**, S208–S217, doi:10.1016/j.eneco.2012.09.011.
21 <http://dx.doi.org/10.1016/j.eneco.2012.09.011>.
- 22 Boillat, S., and F. Berkes, 2013: Perception and interpretation of climate change among quechua
23 farmers of bolivia: Indigenous knowledge as a resource for adaptive capacity. *Ecol. Soc.*,
24 doi:10.5751/ES-05894-180421.
- 25 Bommarco, R., D. Kleijn, and S. G. Potts, 2013: Ecological intensification: Harnessing ecosystem
26 services for food security. *Trends Ecol. Evol.*, doi:10.1016/j.tree.2012.10.012.
- 27 Bommel, P., and Coauthors, 2014: A Further Step Towards Participatory Modelling. Fostering
28 Stakeholder Involvement in Designing Models by Using Executable UML. *J. Artif. Soc. Soc.*
29 *Simul.*, **17**, doi:10.18564/jasss.2381. <http://jasss.soc.surrey.ac.uk/17/1/6.html> (Accessed
30 December 12, 2017).
- 31 Bonan, G. B., 2008: Forests and climate change: Forcings, feedbacks, and the climate benefits of
32 forests. *Science (80-.)*, **320**, 1444–1449, doi:10.1126/science.1155121.
- 33 Bonatti, M., and Coauthors, 2016: Climate vulnerability and contrasting climate perceptions as an
34 element for the development of community adaptation strategies: Case studies in Southern
35 Brazil. *Land use policy*, doi:10.1016/j.landusepol.2016.06.033.
- 36 Bonney, R., J. L. Shirk, T. B. Phillips, A. Wiggins, H. L. Ballard, A. J. Miller-Rushing, and J. K.
37 Parrish, 2014: Citizen science. Next steps for citizen science. *Science*, **343**, 1436–1437,
38 doi:10.1126/science.1251554. <http://www.ncbi.nlm.nih.gov/pubmed/24675940> (Accessed
39 December 12, 2017).
- 40 Bonsch, M., and Coauthors, 2016: Trade-offs between land and water requirements for large-scale
41 bioenergy production. *GCB Bioenergy*, **8**, 11–24, doi:10.1111/gcbb.12226.
42 <http://doi.wiley.com/10.1111/gcbb.12226> (Accessed December 9, 2017).
- 43 Bonsu, N. O., Á. N. Dhubháin, and D. O’Connor, 2017: Evaluating the use of an integrated forest
44 land-use planning approach in addressing forest ecosystem services conflicting demands:
45 Expreince within an Irish forest landscape. *Futures*, **86**, 1–17,
46 doi:10.1016/j.futures.2016.08.004.
- 47 Boonstra, W. J., and T. T. H. Hanh, 2015: Adaptation to climate change as social–ecological trap: a
48 case study of fishing and aquaculture in the Tam Giang Lagoon, Vietnam. *Environ. Dev.*

- 1 *Sustain.*, **17**, 1527–1544.
- 2 Borras Jr., S. M., R. Hall, I. Scoones, B. White, and W. Wolford, 2011: Towards a Better
3 Understanding of Global Land Grabbing: an Editorial Introduction. *J. Peasant Stud.*, **38**, 209–
4 216, doi:10.1080/03066150.2011.559005.
- 5 Borras Jr, S. M., and Coauthors, 2011: Towards a better understanding of global land grabbing: An
6 editorial introduction. *J. Peasant Stud.*, **38**, 209–216, doi:10.1080/03066150.2011.559005.
- 7 Borychowski, M., and A. Czyżewski, 2015: Determinants of prices increase of agricultural
8 commodities in a global context1. *Management*, doi:10.1515/manment-2015-0020.
- 9 Boyce, J. K., 2018: Carbon Pricing: Effectiveness and Equity. *Ecol. Econ.*, **150**, 52–61,
10 doi:<https://doi.org/10.1016/j.ecolecon.2018.03.030>.
11 <http://www.sciencedirect.com/science/article/pii/S092180091731580X>.
- 12 Boyd, E., and C. Folke, 2011a: *Adapting institutions: Governance, complexity and social–ecological*
13 *resilience*. 1-290 pp.
- 14 —, and —, 2011b: *Adapting institutions: Governance, complexity and social–ecological*
15 *resilience*.
- 16 —, and —, 2012: Adapting Institutions, Adaptive Governance and Complexity: an introduction.
17 *Adapting Institutions: Governance, complexity and social-ecological resilience*, E. Boyd and C.
18 Folke, Eds., Cambridge UP, Cambridge, United Kingdom and New York, NY, USA, 1–8.
- 19 —, B. Nykvist, S. Borgström, and I. A. Stacewicz, 2015: Anticipatory governance for social-
20 ecological resilience. *Ambio*, **44**, 149–161, doi:10.1007/s13280-014-0604-x.
- 21 Boyle, Michelle; Kay, James J.; Pond, B., 2001: MONITORING IN SUPPORT OF POLICY: AN
22 ADAPTIVE ECOSYSTEM APPROACH. *Encyclopedia Glob. Environ. Chang.*, **4**, 116–137.
- 23 Boysen, L. R., W. Lucht, D. Gerten, and V. Heck, 2016: Impacts devalue the potential of large-scale
24 terrestrial CO₂ removal through biomass plantations. *Environ. Res. Lett.*, **11**, 095010,
25 doi:10.1088/1748-9326/11/9/095010. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=9/a=095010?key=crossref.244d75c1a47b210f79f0a1cb65588a46)
26 9326/11/i=9/a=095010?key=crossref.244d75c1a47b210f79f0a1cb65588a46 (Accessed
27 December 9, 2017).
- 28 Boysen, L. R., W. Lucht, and D. Gerten, 2017a: Trade-offs for food production, nature conservation
29 and climate limit the terrestrial carbon dioxide removal potential. *Glob. Chang. Biol.*, **23**, 4303–
30 4317, doi:10.1111/gcb.13745.
- 31 —, —, —, V. Heck, T. M. Lenton, and H. J. Schellnhuber, 2017b: The limits to global-
32 warming mitigation by terrestrial carbon removal. *Earth’s Futur.*, **5**, 463–474,
33 doi:10.1002/2016EF000469.
- 34 Brancalion, P. H. S., R. A. G. Viani, B. B. N. Strassburg, and R. R. Rodrigues, 2012: Finding the
35 money for tropical forest restoration. *Unasylva*, **63**, 239.
- 36 Brander, K., and Keith, 2015: Improving the Reliability of Fishery Predictions Under Climate
37 Change. *Curr. Clim. Chang. Reports*, **1**, 40–48, doi:10.1007/s40641-015-0005-7.
- 38 Brander, K. M., 2007: Global fish production and climate change. *Proc. Natl. Acad. Sci.*, **104**, 19709–
39 19714.
- 40 Branger, F., and P. Quirion, 2014: Climate policy and the ‘ carbon haven ’ effect. *Wiley Interdiscip.*
41 *Rev. Clim. Chang.*, **5**, doi:10.1002/wcc.245.
- 42 Bratton, M., 2007: Formal versus informal institutions in Africa. *J. Democr.*, **18**, 96–110,
43 doi:10.1353/jod.2007.0041.
- 44 Brechin, S. R., and M. I. Espinoza, 2017: A case for further refinement of the green climate fund’s
45 50:50 ratio climate change mitigation and adaptation allocation framework: Toward a more
46 targeted approach. *Clim. Change*, **142**, 311–320, doi:10.1007/s10584-017-1938-8.

- 1 Breininger, D., B. Duncan, M. Eaton, F. Johnson, and J. Nichols, 2014: Integrating Land Cover
2 Modeling and Adaptive Management to Conserve Endangered Species and Reduce Catastrophic
3 Fire Risk. *Land*, **3**, 874–897, doi:10.3390/land3030874.
- 4 Brenkert–Smith, H., P. A. Champ, and N. Flores, 2006: Insights into wildfire mitigation decisions
5 among wildland–urban interface residents. *Soc. Nat. Resour.*, **19**, 759–768.
- 6 Bresch, D. N., and Coauthors, 2017: Sovereign Climate and Disaster Risk Pooling. *World Bank Tech.*
7 *Contrib. to G20*,. <https://www.research-collection.ethz.ch/handle/20.500.11850/219554>
8 (Accessed May 25, 2018).
- 9 Bright, R. M., E. Davin, T. O’Halloran, J. Pongratz, K. Zhao, and A. Cescatti, 2017: Local
10 temperature response to land cover and management change driven by non-radiative processes.
11 *Nat. Clim. Chang.*, **7**, doi:10.1038/nclimate3250.
- 12 Broegaard, R. B., T. Vongvisouk, and O. Mertz, 2017: Contradictory Land Use Plans and Policies in
13 Laos: Tenure Security and the Threat of Exclusion. *World Dev.*, **89**, 170–183,
14 doi:10.1016/J.WORLDDEV.2016.08.008.
- 15 Brondizio, E. S., and F.-M. L. Tourneau, 2016: Environmental governance for all. *Science (80-.)*,
16 **352**, 1272–1273, doi:10.1126/science.aaf5122.
17 <http://www.sciencemag.org/cgi/doi/10.1126/science.aaf5122> (Accessed April 15, 2019).
- 18 Bronen, R., and F. S. Chapin, 2013: Adaptive governance and institutional strategies for climate-
19 induced community relocations in Alaska. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1210508110.
- 20 Brooks, J., 2014: Policy coherence and food security: The effects of OECD countries’ agricultural
21 policies. *Food Policy*, doi:10.1016/j.foodpol.2013.10.006.
- 22 Brown, C., P. Alexander, S. Holzhauer, and M. D. A. Rounsevell, 2017: Behavioral models of climate
23 change adaptation and mitigation in land-based sectors. *Wiley Interdiscip. Rev. Clim. Chang.*, **8**,
24 e448, doi:10.1002/wcc.448. <http://doi.wiley.com/10.1002/wcc.448> (Accessed December 12,
25 2017).
- 26 Brown, I., and M. Castellazzi, 2014: Scenario analysis for regional decision-making on sustainable
27 multifunctional land uses. *Reg. Environ. Chang.*, **14**, 1357–1371, doi:10.1007/s10113-013-0579-
28 3.
- 29 Brown, M. L., 2010: Limiting Corrupt Incentives in a Global REDD Regime. *Ecol. Law Q.*, **37**, 237–
30 267, doi:10.15779/Z38HC41. <http://dx.doi.org/https://doi.org/10.15779/Z38HC41> (Accessed
31 May 23, 2018).
- 32 Bruce, J. W., and S. E. Migot-Adholla, 1994: Introduction: Are Indigenous African Tenure Systems
33 Insecure? *Searching for Land Tenure Security in Africa*, Ohn W Bruce and Shem E Migot-
34 Adholla, Ed., Kendall/Hunt, p. 282
35 [http://www.cabdirect.org/abstracts/19951812654.html;jsessionid=62E08AE916A2DB0B48D97
36 616CFD46ACD](http://www.cabdirect.org/abstracts/19951812654.html;jsessionid=62E08AE916A2DB0B48D97616CFD46ACD) (Accessed May 23, 2018).
- 37 de Bruin, K., R. Dellink, S. Agrawala, and R. Dellink, 2009: Economic Aspects of Adaptation to
38 Climate Change: Integrated Assessment Modelling of Adaptation Costs and Benefits. *OECD*
39 *Environ. Work. Pap.*, 22; 36–38, doi:10.1787/225282538105.
- 40 Bruvoll, A., and B. M. Larsen, 2004: Greenhouse gas emissions in Norway: Do carbon taxes work?
41 *Energy Policy*, **32**, 493–505, doi:10.1016/S0301-4215(03)00151-4.
- 42 Bryan, B. A., D. King, and E. Wang, 2010: Biofuels agriculture: landscape-scale trade-offs between
43 fuel, economics, carbon, energy, food, and fiber. *Gcb Bioenergy*, **2**, 330–345.
- 44 Bryngelsson, D., S. Wirsenius, F. Hedenus, and U. Sonesson, 2016: How can the EU climate targets
45 be met? A combined analysis of technological and demand-side changes in food and agriculture.
46 *Food Policy*, doi:10.1016/j.foodpol.2015.12.012.
- 47 Burby, R. J., and P. J. May, 2009: *Command or Cooperate? Rethinking Traditional Central*

- 1 *Governments' Hazard Mitigation Policies*. 21-33 pp. <http://ebooks.iospress.nl/publication/24857>
2 (Accessed February 26, 2019).
- 3 Burton, R. J. F., C. Kuczera, and G. Schwarz, 2008: Exploring farmers' cultural resistance to
4 voluntary agri-environmental schemes. *Sociol. Ruralis*, **48**.
- 5 Busch, J., 2018: Monitoring and evaluating the payment-for-performance premise of REDD+: the
6 case of India's ecological fiscal transfers. *Ecosyst. Heal. Sustain.*, **4**, 169–175.
- 7 ———, and A. Mukherjee, 2017: Encouraging State Governments to Protect and Restore Forests Using
8 Ecological Fiscal Transfers: India's Tax Revenue Distribution Reform. *Conserv. Lett.*, **00**, 1–10,
9 doi:10.1111/conl.12416. <http://doi.wiley.com/10.1111/conl.12416>.
- 10 Byerlee, D., X. Rueda, D. Byerlee, and X. Rueda, 2015: From Public to Private Standards for Tropical
11 Commodities: A Century of Global Discourse on Land Governance on the Forest Frontier.
12 *Forests*, **6**, 1301–1324, doi:10.3390/f6041301. <http://www.mdpi.com/1999-4907/6/4/1301>
13 (Accessed April 3, 2019).
- 14 Byers, E., and Coauthors, 2018a: Global exposure and vulnerability to multi-sector development and
15 climate change hotspots. *Environ. Res. Lett.*, **13**, 055012, doi:10.1088/1748-9326/aabf45.
16 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=5/a=055012?key=crossref.dcb006b2e0b98d78e8d8ed4aa6eb51fb)
17 [9326/13/i=5/a=055012?key=crossref.dcb006b2e0b98d78e8d8ed4aa6eb51fb](http://stacks.iop.org/1748-9326/13/i=5/a=055012?key=crossref.dcb006b2e0b98d78e8d8ed4aa6eb51fb) (Accessed April 11,
18 2019).
- 19 ———, and Coauthors, 2018b: Global exposure and vulnerability to multi-sector development and
20 climate change hotspots. *Environ. Res. Lett.*, **13**, 055012, doi:10.1088/1748-9326/aabf45.
- 21 Camacho, A. E., 2009: Adapting Governance to Climate Change: Managing Uncertainty through a
22 Learning Infrastructure. *Emory Law J.*, doi:10.2139/ssrn.1352693.
- 23 Camacho, L. D., D. T. Gevaña, †Antonio P. Carandang, and S. C. Camacho, 2016: Indigenous
24 knowledge and practices for the sustainable management of Ifugao forests in Cordillera,
25 Philippines. *Int. J. Biodivers. Sci. Ecosyst. Serv. Manag.*, **12**, 5–13,
26 doi:10.1080/21513732.2015.1124453.
27 <https://www.tandfonline.com/doi/full/10.1080/21513732.2015.1124453> (Accessed April 23,
28 2019).
- 29 Cameron, C., S. Pachauri, N. D. Rao, D. McCollum, J. Rogelj, and K. Riahi, 2016: Policy trade-offs
30 between climate mitigation and clean cook-stove access in South Asia. *Nat. Energy*, **1**,
31 doi:10.1038/nenergy.2015.10.
32 <https://pdfs.semanticscholar.org/b5a1/a739df5dfcff81cf5c989ed9bc2050c5eba2.pdf> (Accessed
33 May 18, 2018).
- 34 Campillo, G., M. Mullan, and L. Vallejo, 2017: Climate Change Adaptation and Financial Protection.
35 doi:10.1787/0b3dc22a-en. [http://www.oecd-ilibrary.org/environment/climate-change-](http://www.oecd-ilibrary.org/environment/climate-change-adaptation-and-financial-protection_0b3dc22a-en)
36 [adaptation-and-financial-protection_0b3dc22a-en](http://www.oecd-ilibrary.org/environment/climate-change-adaptation-and-financial-protection_0b3dc22a-en) (Accessed May 25, 2018).
- 37 Caney, S., 2014: Climate change, intergenerational equity and the social discount rate. *Polit. Philos.*
38 *Econ.*, doi:10.1177/1470594X14542566.
- 39 Canonico, G. C., A. Arthington, J. K. McCrary, and M. L. Thieme, 2005: The effects of introduced
40 tilapias on native biodiversity. *Aquat. Conserv. Mar. Freshw. Ecosyst.*, **15**, 463–483.
- 41 Cantarello, E., C. A. Aquino, J. C. Birch, C. Echeverria, N. T. Garavito, T. Kitzberger, A. C. Newton,
42 and I. Schiappacasse, 2010: Cost-effectiveness of dryland forest restoration evaluated by spatial
43 analysis of ecosystem services. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1003369107.
- 44 Capstick, S., L. Whitmarsh, W. Poortinga, N. Pidgeon, and P. Upham, 2015: International trends in
45 public perceptions of climate change over the past quarter century. *Wiley Interdiscip. Rev. Clim.*
46 *Chang.*, **6**, 35–61, doi:10.1002/wcc.321. <http://doi.wiley.com/10.1002/wcc.321> (Accessed
47 October 2, 2018).
- 48 Cardona, O., 2012: Determinants of Risk : Exposure and Vulnerability Coordinating Lead Authors :

- 1 Lead Authors : Review Editors : Contributing. *Manag. Risks Extrem. Events Disasters to Adv.*
2 *Clim. Chang. Adapt. - A Spec. Rep. Work. Groups I II Intergov. Panel Clim. Chang.*,
3 doi:10.1017/CBO9781139177245.005.
- 4 Carleton, T. A., and S. M. Hsiang, 2016a: Social and economic impacts of climate. *Science (80-.)*,
5 **353**, doi:10.1126/science.aad9837.
- 6 —, and —, 2016b: Social and economic impacts of climate. *Science (80-.)*,
7 doi:10.1126/science.aad9837.
- 8 Carroll, R., A. Stern, D. Zook, R. Funes, A. Rastegar, and Y. Lien, 2012: *Catalyzing Smallholder*
9 *AGricultural Finance*.
- 10 Carter, A., 2017: Placeholders and Changemakers: Women Farmland Owners Navigating Gendered
11 Expectations. *Rural Sociol.*, **82**, 499–523, doi:10.1111/ruso.12131.
- 12 Carter, E., M. Shan, Y. Zhong, W. Ding, Y. Zhang, J. Baumgartner, and X. Yang, 2018: Development
13 of renewable, densified biomass for household energy in China. *Energy Sustain. Dev.*, **46**, 42–
14 52, doi:10.1016/j.esd.2018.06.004. <https://doi.org/10.1016/j.esd.2018.06.004> (Accessed April
15 15, 2019).
- 16 Carter, S., A. M. Manceur, R. Seppelt, K. Hermans-Neumann, M. Herold, and L. Verchot, 2017:
17 Large scale land acquisitions and REDD plus : a synthesis of conflictsand opportunities.
18 *Environ. Res. Lett.*, **12**, doi:10.1088/1748-9326/aa6056.
- 19 Carvalho, J. L. N., T. W. Hudiburg, H. C. J. Franco, and E. H. Delucia, 2016: Contribution of above-
20 and belowground bioenergy crop residues to soil carbon. *GCB Bioenergy*,
21 doi:10.1111/gcbb.12411.
- 22 Casellas Connors, J. P., and A. Janetos, 2016: Assessing the Impacts of Multiple Breadbasket
23 Failures. *AGU Fall Meeting Abstracts*.
- 24 Cash, D. W., W. N. Adger, F. Berkes, P. Garden, L. Lebel, P. Olsson, L. Pritchard, and O. Young,
25 2006: Scale and Cross-Scale Dynamics: Governance and Information in a Multilevel World.
26 *Ecol. Soc.*, **11**, art8, doi:10.5751/ES-01759-110208.
27 <http://www.ecologyandsociety.org/vol11/iss2/art8/> (Accessed December 21, 2017).
- 28 Castán Broto, V., 2017: Urban Governance and the Politics of Climate change. *World Dev.*,
29 doi:10.1016/j.worlddev.2016.12.031.
- 30 Castella, J.-C., J. Bourgoin, G. Lestrelin, and B. Bouahom, 2014: A model of the science–practice–
31 policy interface in participatory land-use planning: lessons from Laos. *Landsc. Ecol.*, **29**, 1095–
32 1107, doi:10.1007/s10980-014-0043-x. <http://link.springer.com/10.1007/s10980-014-0043-x>
33 (Accessed December 12, 2017).
- 34 Cavicchi, B., S. Palmieri, M. Odaldi, B. Cavicchi, S. Palmieri, and M. Odaldi, 2017: The Influence of
35 Local Governance: Effects on the Sustainability of Bioenergy Innovation. *Sustainability*, **9**, 406,
36 doi:10.3390/su9030406. <http://www.mdpi.com/2071-1050/9/3/406> (Accessed April 11, 2019).
- 37 CDM Executive Board, 2012: *Default values of fraction of non-renewable biomass for least*
38 *developed countries and small island developing States*. 13 pp.
39 https://cdm.unfccc.int/Reference/Notes/meth/meth_note12.pdf (Accessed April 14, 2019).
- 40 Ceddia, M. ., U. Gunter, and A. Corriveau-Bourque, 2015: Land tenure and agricultural expansion in
41 Latin America: The role of Indigenous Peoples’ and local communities’ forest rights. *Glob.*
42 *Environ. Chang.*, **35**, 316–322, doi:10.1016/j.gloenvcha.2015.09.010.
- 43 Celentano, D., G. X. Rousseau, V. L. Engel, M. Zelarayán, E. C. Oliveira, A. C. M. Araujo, and E. G.
44 de Moura, 2017: Degradation of Riparian Forest Affects Soil Properties and Ecosystem Services
45 Provision in Eastern Amazon of Brazil. *L. Degrad. Dev.*, doi:10.1002/ldr.2547.
- 46 Cerutti, P. O., and Coauthors, 2015: The socioeconomic and environmental impacts of wood energy
47 value chains in Sub-Saharan Africa: a systematic map protocol. *Environ. Evid.*, **4**, 12,

- 1 doi:10.1186/s13750-015-0038-3.
2 <http://environmentalevidencejournal.biomedcentral.com/articles/10.1186/s13750-015-0038-3>
3 (Accessed November 16, 2018).
- 4 Chadburn, S. E., Burke, E. J., Cox, P. M., Friedlingstein, P., Hugelius, G., & Westermann, S. (2017).
5 An observation-based constraint on permafrost loss as a function of global warming. *Nature*
6 *Climate Change*. <https://doi.org/10.1038/nclimate3262>
- 7 Chaffin, B. C., H. Gosnell, and B. A. Cosens, 2014: A decade of adaptive governance scholarship:
8 Synthesis and future directions. *Ecol. Soc.*, doi:10.5751/ES-06824-190356.
- 9 Challinor, A. J., W. N. Adger, and T. G. Benton, 2017: Climate risks across borders and scales. *Nat.*
10 *Clim. Chang.*, **7**, 621–623, doi:10.1038/nclimate3380.
- 11 Chambers, R., 1988: *Managing canal irrigation: practical analysis from South Asia*. Cambridge
12 University Press, 279 pp. https://books.google.co.uk/books?hl=en&lr=&id=matcBqmG8C&oi=fnd&pg=PR13&dq=chambers+managing+canal+irrigation&ots=_xx9DbmVbQ&sig=NLHWRJYCCnXekZt1sqzj2Qhb2c0#v=onepage&q=chambers+managing+canal+irrigation&f=false (Accessed April 12, 2019).
- 16 Chambwera, M., and G. Heal, 2014: Economics of Adaptation. *Climate Change 2014 Impacts,*
17 *Adaptation and Vulnerability: Part A: Global and Sectoral Aspects*, 945–977.
- 18 Chambwera, M., G. Heal, C. Dubeux, S. Hallegatte, L. Leclerc, A. Markandya, B. A. McCarl, and R.
19 Mechler, 2014a: Economics of adaptation. *Impacts, Adaptation, and Vulnerability. Part A:*
20 *Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report*
21 *of the IPCC*, 945–977.
- 22 —, —, —, —, —, —, —, and —, 2014b: *Economics of adaptation*.
- 23 Chambwera, M., and Coauthors, 2015: Economics of adaptation. *Climate Change 2014 Impacts,*
24 *Adaptation and Vulnerability: Part A: Global and Sectoral Aspects*.
- 25 Chandra, A., P. P. Saradhi, R. K. Maikhuri, K. G. Saxena, and K. S. Rao, 2011: Traditional
26 agrodiversity management: A case study of central himalayan village ecosystem. *J. Mt. Sci.*, **8**,
27 62–74, doi:10.1007/s11629-011-1081-3. <http://link.springer.com/10.1007/s11629-011-1081-3>
28 (Accessed April 6, 2019).
- 29 Chaney, P., and R. Fevre, 2001: Inclusive governance and “minority” groups: The role of the third
30 sector in Wales. *Voluntas*, **12**, 131–156, doi:10.1023/A:1011286602556.
- 31 Chang, S. E., T. L. McDaniels, J. Mikawoz, and K. Peterson, 2007: Infrastructure failure
32 interdependencies in extreme events: power outage consequences in the 1998 Ice Storm. *Nat.*
33 *Hazards*, **41**, 337–358.
- 34 Chang, Y., G. Li, Y. Yao, L. Zhang, and C. Yu, 2016: Quantifying the water-energy-food nexus:
35 Current status and trends. *Energies*, doi:10.3390/en9020065.
- 36 Chanza, N., and A. de Wit, 2016: Enhancing climate governance through indigenous knowledge:
37 Case in sustainability science. *S. Afr. J. Sci.*, **Volume 112**, 1–7,
38 doi:10.17159/sajs.2016/20140286. <http://sajs.co.za/article/view/4058> (Accessed April 9, 2019).
- 39 Chapin, F. S., and Coauthors, 2010: Resilience of Alaska’s boreal forest to climatic change. *Can. J.*
40 *For. Res.*, **40**, 1360–1370.
- 41 Chasek, P., U. Safriel, S. Shikongo, and V. F. Fuhrman, 2015: Operationalizing Zero Net Land
42 Degradation: The next stage in international efforts to combat desertification? *J. Arid Environ.*,
43 **112**, 5–13, doi:10.1016/j.jaridenv.2014.05.020.
- 44 Chatzopoulos, T., I. Pérez Domínguez, M. Zampieri, and A. Toreti, 2019: Climate extremes and
45 agricultural commodity markets: A global economic analysis of regionally simulated events.
46 *Weather Clim. Extrem.*, doi:10.1016/j.wace.2019.100193.
- 47 Chávez Michaelsen, A., S. G. Perz, L. Huamani Briceño, R. Fernandez Menis, N. Bejar Chura, R.

- 1 Moreno Santillan, J. Díaz Salinas, and I. F. Brown, 2017: Effects of drought on deforestation
2 estimates from different classification methodologies: Implications for REDD+ and other
3 payments for environmental services programs. *Remote Sens. Appl. Soc. Environ.*, **5**, 36–44,
4 doi:10.1016/J.RSASE.2017.01.003.
5 <https://www.sciencedirect.com/science/article/pii/S2352938516300714> (Accessed October 15,
6 2018).
- 7 Chen, D., W. Wei, and L. Chen, 2017a: Effects of terracing practices on water erosion control in
8 China: A meta-analysis. *Earth-Science Rev.*, **173**, 109–121,
9 doi:10.1016/J.EARSCIREV.2017.08.007.
10 <https://www.sciencedirect.com/science/article/pii/S0012825217300090> (Accessed April 8,
11 2019).
- 12 Chen, J., 2011: MODERN DISASTER THEORY : EVALUATING DISASTER LAW AS A
13 PORTFOLIO OF LEGAL RULES. *Emory Int Law Rev.*,
- 14 Chen, Y., Y. Tan, and Y. Luo, 2017b: Post-disaster resettlement and livelihood vulnerability in rural
15 China. *Disaster Prev. Manag.*, doi:10.1108/DPM-07-2016-0130.
- 16 Cheng, C.-Y., and J. Urpelainen, 2014: Fuel stacking in India: Changes in the cooking and lighting
17 mix, 1987e2010 *. *Energy*, 306–317, doi:10.1016/j.energy.2014.08.023.
18 <http://dx.doi.org/10.1016/j.energy.2014.08.023> (Accessed April 14, 2019).
- 19 Chhatre, A., and A. Agrawal, 2009: Trade-offs and synergies between carbon storage and livelihood
20 benefits from forest commons. *Proc. Natl. Acad. Sci.*, **106**, 17667–17670.
- 21 —, and S. Lakhanpal, 2018: For the environment, against conservation: Conflict between
22 renewable energy and biodiversity protection in India. *Conservation and Development in India*,
23 Routledge, 52–72.
- 24 Chomba, S., J. Kariuki, J. F. Lund, and F. Sinclair, 2016: Roots of inequity: How the implementation
25 of REDD+ reinforces past injustices. *Land use policy*, **50**, 202–213,
26 doi:10.1016/j.landusepol.2015.09.021.
- 27 Chopra, K., 2017: Land and Forest Policy: Resources for Development or Our Natural Resources?
28 *Development and Environmental Policy in India*, Springer, 13–25.
- 29 Christenson, E., M. Elliott, O. Banerjee, L. Hamrick, and J. Bartram, 2014: Climate-related hazards:
30 A method for global assessment of urban and rural population exposure to cyclones, droughts,
31 and floods. *Int. J. Environ. Res. Public Health*, **11**, 2169–2192, doi:10.3390/ijerph110202169.
- 32 Christian-Smith, J., M. C. Levy, and P. H. Gleick, 2015: Maladaptation to drought: a case report from
33 California, USA. *Sustain. Sci.*, **10**, 491–501, doi:10.1007/s11625-014-0269-1.
34 <https://doi.org/10.1007/s11625-014-0269-1>.
- 35 Chung Tiam Fook, T., 2017: Transformational processes for community-focused adaptation and
36 social change: a synthesis. *Clim. Dev.*, doi:10.1080/17565529.2015.1086294.
- 37 Church, S. P., and Coauthors, 2017: Agricultural trade publications and the 2012 Midwestern U.S.
38 drought: A missed opportunity for climate risk communication. *Clim. Risk Manag.*,
39 doi:10.1016/j.crm.2016.10.006.
- 40 Clarke, D., and S. Dercon, 2016a: *Dull Disasters? How Planning Ahead Will Make a Difference*.
41 Oxford University Press, Oxford,.
- 42 —, and —, 2016b: *Dull Disasters? How Planning Ahead Will Make a Difference*. First. Oxford
43 University Press, Oxford,.
- 44 Clarke, D. J., 2016: A Theory of Rational Demand for Index Insurance. *Am. Econ. J.*
45 *Microeconomics*, **8**, 283–306, doi:10.1257/mic.20140103.
46 <http://pubs.aeaweb.org/doi/10.1257/mic.20140103> (Accessed April 3, 2019).
- 47 Clarke, J., and M. Alston, 2017: Understanding the "local" and "global":

- 1 intersections engendering change for women in family farming in Australia | Request PDF.
2 *Women in Agriculture Worldwide: Key issues and practical approaches*, A. Fletcher and W.
3 Kubik, Eds., Routledge, Abingdon and New York, 13–22
4 https://www.researchgate.net/publication/308893180_Understanding_the_local_and_global_inte
5 [rsections_engendering_change_for_women_in_family_farming_in_Australia](https://www.researchgate.net/publication/308893180_Understanding_the_local_and_global_inte) (Accessed April 4,
6 2019).
- 7 Clarvis, M. H., E. Bohensky, and M. Yarime, 2015: Can resilience thinking inform resilience
8 investments? Learning from resilience principles for disaster risk reduction. *Sustain.*, **7**, 9048–
9 9066, doi:10.3390/su7079048.
- 10 Clemens, H., R. Bailis, A. Nyambane, and V. Ndung'u D, 2018: Africa Biogas Partnership Program:
11 A review of clean cooking implementation through market development in East Africa. *Energy*
12 *Sustain. Dev.*, **46**, 23–31, doi:10.1016/j.esd.2018.05.012.
13 <https://doi.org/10.1016/j.esd.2018.05.012> (Accessed April 15, 2019).
- 14 Clemens, M., J. Rijke, A. Pathirana, J. Evers, and N. Hong Quan, 2015: Social learning for adaptation
15 to climate change in developing countries: insights from Vietnam. *J. Water Clim. Chang.*, **7**,
16 jwc2015004, doi:10.2166/wcc.2015.004. <https://iwaponline.com/jwcc/article/7/2/365-378/416>
17 (Accessed October 2, 2018).
- 18 Clover, J., and S. Eriksen, 2009: The effects of land tenure change on sustainability: human security
19 and environmental change in southern African savannas. *Environ. Sci. Policy*, **12**, 53–70,
20 doi:10.1016/j.envsci.2008.10.012.
- 21 Coady, D., I. Parry, L. Sears, and B. Shang, 2017: How Large Are Global Fossil Fuel Subsidies?
22 *World Dev.*, doi:10.1016/j.worlddev.2016.10.004.
- 23 Coates, J., 2013: Build it back better: Deconstructing food security for improved measurement and
24 action. *Glob. Food Sec.*, doi:10.1016/j.gfs.2013.05.002.
- 25 Cochran, F. V., N. A. Brunzell, A. Cabalzar, P.-J. van der Veld, E. Azevedo, R. A. Azevedo, R. A.
26 Pedrosa, and L. J. Winegar, 2016: Indigenous ecological calendars define scales for climate
27 change and sustainability assessments. *Sustain. Sci.*, **11**, 69–89, doi:10.1007/s11625-015-0303-y.
28 <http://link.springer.com/10.1007/s11625-015-0303-y> (Accessed April 6, 2019).
- 29 Coenen, F., and F. H. J. M. Coenen, 2009: *Public Participation and Better Environmental Decisions*.
30 183–209 pp.
- 31 Cohen, R., and M. Bradley, 2010: Disasters and Displacement: Gaps in Protection. *J. Int. Humanit.*
32 *Leg. Stud.*, doi:10.1163/187815210X12766020139884.
- 33 Cole, S., 2015: Overcoming Barriers to Microinsurance Adoption: Evidence from the Field†. *Geneva*
34 *Pap. Risk Insur. - Issues Pract.*, **40**, 720–740.
35 <https://ideas.repec.org/a/pal/gpprii/v40y2015i4p720-740.html> (Accessed April 3, 2019).
- 36 ———, X. Giné, J. Tobacman, P. Topalova, R. Townsend, and J. Vickery, 2013: Barriers to Household
37 Risk Management: Evidence from India. *Am. Econ. J. Appl. Econ.*, **5**, 104–135,
38 doi:10.1257/app.5.1.104. <http://pubs.aeaweb.org/doi/10.1257/app.5.1.104> (Accessed April 3,
39 2019).
- 40 Collar, N. J., P. Patil, and G. S. Bhardwaj, 2015: What can save the Great Indian Bustard *Ardeotis*
41 *nigriceps*. *Bird. ASIA*, **23**, 15–24.
- 42 Collins, A. M., 2014: Governing the Global Land Grab: What role for Gender in the
43 Voluntary Guidelines and the Principles for Responsible Investment? *GLOBALIZATIONS*, **11**,
44 189–203, doi:10.1080/14747731.2014.887388.
- 45 Collins, R. D., R. de Neufville, J. Claro, T. Oliveira, and A. P. Pacheco, 2013: Forest fire management
46 to avoid unintended consequences: A case study of Portugal using system dynamics. *J. Environ.*
47 *Manage.*, **130**, 1–9, doi:10.1016/j.jenvman.2013.08.033.
- 48 de Coninck, H., and Coauthors, 2018: Strengthening and implementing the global response. *Global*

- 1 *Warming of 1.5C: an IPCC special report on the impacts of global warming of 1.5C above pre-*
2 *industrial levels and related global greenhouse gas emission pathways, in the context of*
3 *strengthening the global response to the threat of climate change*
4 <http://www.ipcc.ch/report/sr15/>.
- 5 Conrad, C. C., and K. G. Hilchey, 2011: A review of citizen science and community-based
6 environmental monitoring: issues and opportunities. *Environ. Monit. Assess.*, **176**, 273–291,
7 doi:10.1007/s10661-010-1582-5. <http://link.springer.com/10.1007/s10661-010-1582-5>
8 (Accessed December 11, 2017).
- 9 Conradt, S., R. Finger, and M. Spörri, 2015: Flexible weather index-based insurance design. *Clim.*
10 *Risk Manag.*, **10**, 106–117, doi:10.1016/j.crm.2015.06.003.
- 11 Conway, D., and E. L. F. Schipper, 2011: Adaptation to climate change in Africa: Challenges and
12 opportunities identified from Ethiopia. *Glob. Environ. Chang.*, **21**, 227–237,
13 doi:10.1016/j.gloenvcha.2010.07.013.
- 14 ———, and Coauthors, 2015: Climate and southern Africa’s water–energy–food nexus. *Nat. Clim.*
15 *Chang.*, **5**, 837. <http://dx.doi.org/10.1038/nclimate2735>.
- 16 Cook, S., and J. Pincus, 2015: Poverty, Inequality and Social Protection in Southeast Asia: An
17 Introduction. *Southeast Asian Econ.*, doi:10.1355/ae31-1a.
- 18 Cooke, S. J., and Coauthors, 2016: On the sustainability of inland fisheries: finding a future for the
19 forgotten. *Ambio*, **45**, 753–764.
- 20 Cools, J., D. Innocenti, and S. O’Brien, 2016: Lessons from flood early warning systems. *Environ.*
21 *Sci. Policy*, **58**, 117–122, doi:10.1016/J.ENVSCL.2016.01.006.
22 <https://www.sciencedirect.com/science/article/pii/S1462901116300065> (Accessed February 25,
23 2019).
- 24 Cooper, M. H., and C. Rosin, 2014: Absolving the sins of emission: The politics of regulating
25 agricultural greenhouse gas emissions in New Zealand. *J. Rural Stud.*, **36**, 391–400,
26 doi:<http://dx.doi.org/10.1016/j.jrurstud.2014.06.008>.
- 27 Cooper, M. H., J. Boston, and J. Bright, 2013: Policy challenges for livestock emissions abatement:
28 Lessons from New Zealand. *Clim. Policy*, **13**, 110–133, doi:10.1080/14693062.2012.699786.
- 29 Corfee-Morlot, J., L. Kamal-Chaoui, M. Donovan, I. Cochran, A. Robert, and P. Teasdale, 2009:
30 *Cities, Climate Change and Multilevel Governance*. 0-125 pp.
- 31 Corradini, M., V. Costantini, A. Markandya, E. Paglialonga, and G. Sforza, 2018: A dynamic
32 assessment of instrument interaction and timing alternatives in the EU low-carbon policy mix
33 design. *Energy Policy*, doi:10.1016/j.enpol.2018.04.068.
- 34 Corsi, S., L. V. Marchisio, and L. Orsi, 2017: Connecting smallholder farmers to local markets:
35 Drivers of collective action, land tenure and food security in East Chad. *Land use policy*, **68**, 39–
36 47, doi:10.1016/J.LANDUSEPOL.2017.07.025.
37 <https://www.sciencedirect.com/science/article/pii/S0264837716307104> (Accessed May 23,
38 2018).
- 39 Cosens, B., and Coauthors, 2017a: The role of law in adaptive governance. *Ecol. Soc.*, **22**,
40 doi:10.5751/ES-08731-220130.
- 41 Cosens, B. A., and Coauthors, 2017b: The role of law in adaptive governance. *Ecol. Soc.*, **22**,
42 doi:10.5751/ES-08731-220130.
- 43 Costanza, R., R. De Groot, P. Sutton, S. Van Der Ploeg, S. J. Anderson, I. Kubiszewski, S. Farber,
44 and R. K. Turner, 2014: Changes in the global value of ecosystem services. *Glob. Environ.*
45 *Chang.*, **26**, 152–158, doi:10.1016/j.gloenvcha.2014.04.002.
46 <http://dx.doi.org/10.1016/j.gloenvcha.2014.04.002>.
- 47 Costella, C., C. Jaime, J. Arrighi, E. Coughlan de Perez, P. Suarez, and M. van Aalst, 2017a: Scalable

- 1 and Sustainable: How to Build Anticipatory Capacity into Social Protection Systems. *IDS Bull.*,
2 **48**, 31–46, doi:10.19088/1968-2017.151. <http://bulletin.ids.ac.uk/idsbo/article/view/2885>
3 (Accessed May 25, 2018).
- 4 —, —, —, —, —, and —, 2017b: Scalable and Sustainable: How to Build Anticipatory
5 Capacity into Social Protection Systems. *IDS Bull.*, **48**, doi:10.19088/1968-2017.151.
- 6 Cote, M., and A. J. Nightingale, 2012: Resilience thinking meets social theory: Situating social
7 change in socio-ecological systems (SES) research. *Prog. Hum. Geogr.*,
8 doi:10.1177/0309132511425708.
- 9 Cotula, L., 2006a: *Land and water rights in the Sahel: tenure challenges of improving access to*
10 *water for agriculture*. International Institute for Environment and Development, Drylands
11 Programme, 92 pp.
12 [https://books.google.co.uk/books?hl=en&lr=&id=D36jJpNx5S4C&oi=fnd&pg=PA1&dq=Cotul](https://books.google.co.uk/books?hl=en&lr=&id=D36jJpNx5S4C&oi=fnd&pg=PA1&dq=Cotula+a+sahel&ots=u_mbF2pGfm&sig=GqC6CgG8wTIGr4GWELVL1BiPaJc#v=onepage&q=Cotula+sahel&f=false)
13 [a+sahel&ots=u_mbF2pGfm&sig=GqC6CgG8wTIGr4GWELVL1BiPaJc#v=onepage&q=Cotula](https://books.google.co.uk/books?hl=en&lr=&id=D36jJpNx5S4C&oi=fnd&pg=PA1&dq=Cotula+a+sahel&ots=u_mbF2pGfm&sig=GqC6CgG8wTIGr4GWELVL1BiPaJc#v=onepage&q=Cotula+sahel&f=false)
14 [sahel&f=false](https://books.google.co.uk/books?hl=en&lr=&id=D36jJpNx5S4C&oi=fnd&pg=PA1&dq=Cotula+a+sahel&ots=u_mbF2pGfm&sig=GqC6CgG8wTIGr4GWELVL1BiPaJc#v=onepage&q=Cotula+sahel&f=false) (Accessed April 12, 2019).
- 15 —, 2006b: *Land and water rights in the Sahel: tenure challenges of improving access to water for*
16 *agriculture*. International Institute for Environment and Development, Drylands Programme, 92
17 pp.
- 18 Cotula, L., and Coauthors, 2014: Testing Claims about Large Land Deals in Africa: Findings from a
19 Multi-Country Study. *J. Dev. Stud.*, **50**, 903–925, doi:10.1080/00220388.2014.901501.
20 <http://www.tandfonline.com/doi/abs/10.1080/00220388.2014.901501> (Accessed November 2,
21 2018).
- 22 Couvet, D., and A. C. Prevot, 2015: Citizen-science programs: Towards transformative biodiversity
23 governance. *Environ. Dev.*, **13**, 39–45, doi:10.1016/j.envdev.2014.11.003.
- 24 Cowie, A. L., and Coauthors, 2011: Towards sustainable land management in the drylands: Scientific
25 connections in monitoring and assessing dryland degradation, climate change and biodiversity.
26 *L. Degrad. Dev.*, **22**, 248–260, doi:10.1002/ldr.1086.
- 27 Cowie, A. L., and Coauthors, 2018a: Land in balance: The scientific conceptual framework for Land
28 Degradation Neutrality. *Environ. Sci. Policy*, **79**, 25–35, doi:10.1016/j.envsci.2017.10.011.
- 29 —, and Coauthors, 2018b: Land in balance: The scientific conceptual framework for Land
30 Degradation Neutrality. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2017.10.011.
- 31 Craig, R. K., 2010: “Stationary is dead” - long live transformation: five principles for climate change
32 adaptation law. *Harvard Environ. Law Rev.*, **34**, 9–73, doi:10.2139/ssrn.1357766.
33 http://www.law.harvard.edu/students/orgs/elr/vol34_1/9-74.pdf (Accessed May 18, 2018).
- 34 Creutzig, F., P. Agoston, J. C. Goldschmidt, G. Luderer, G. Nemet, and R. C. Pietzcker, 2017: The
35 underestimated potential of solar energy to mitigate climate change. *Nat. Energy*, **2**, 17140.
- 36 Cummins, J. D., and M. A. Weiss, 2016: Equity Capital, Internal Capital Markets, and Optimal
37 Capital Structure in the US Property-Casualty Insurance Industry. *Annu. Rev. Financ. Econ.*, **8**,
38 121–153, doi:10.1146/annurev-financial-121415-032815.
39 <http://www.annualreviews.org/doi/10.1146/annurev-financial-121415-032815> (Accessed April
40 3, 2019).
- 41 Cundale, K., R. Thomas, J. K. Malava, D. Havens, K. Mortimer, and L. Conteh, 2017: A health
42 intervention or a kitchen appliance? Household costs and benefits of a cleaner burning biomass-
43 fuelled cookstove in Malawi. *Soc. Sci. Med.*, **183**, 1–10, doi:10.1016/j.socscimed.2017.04.017.
44 <http://creativecommons.org/licenses/by/4.0/> (Accessed April 15, 2019).
- 45 Cutter, S., and Coauthors, 2012a: Managing the risks from climate extremes at the local level.
46 *Managing the Risks of Extreme Events and Disasters to Advance Climate Change Adaptation*.
- 47 Cutter, S., B. Osman-Elasha, J. Campbell, S.-M. Cheong, S. McCormick, R. Pulwarty, S. Supratid,
48 and G. Ziervogel, 2012b: Managing the Risks from Climate Extremes at the Local Level.

- 1 *Managing the Risks of Extreme Events and Disasters to Advance Climate Change Adaptation:*
2 *Special Report of the Intergovernmental Panel on Climate Change*, 291–338.
- 3 Cutz, L., O. Masera, D. Santana, and A. P. C. Faaij, 2017a: Switching to efficient technologies in
4 traditional biomass intensive countries: The resultant change in emissions. *Energy*, **126**, 513–
5 526, doi:10.1016/J.ENERGY.2017.03.025. [https://www.sciencedirect-](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0360544217303833)
6 [com.ezp.sub.su.se/science/article/pii/S0360544217303833](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0360544217303833) (Accessed May 23, 2018).
- 7 Cutz, L., O. Masera, D. Santana, and A. P. C. Faaij, 2017b: Switching to efficient technologies in
8 traditional biomass intensive countries: The resultant change in emissions.
9 doi:10.1016/j.energy.2017.03.025. <http://dx.doi.org/10.1016/j.energy.2017.03.025> (Accessed
10 November 16, 2018).
- 11 Czembrowski, P., and J. Kronenberg, 2016: Hedonic pricing and different urban green space types
12 and sizes: Insights into the discussion on valuing ecosystem services. *Landsc. Urban Plan.*,
13 doi:10.1016/j.landurbplan.2015.10.005.
- 14 Dale, V. H., R. A. Efroymson, K. L. Kline, and M. S. Davitt, 2015: A framework for selecting
15 indicators of bioenergy sustainability. *Biofuels, Bioprod. Biorefining*, **9**, 435–446,
16 doi:10.1002/bbb.1562.
- 17 Dalin, C., Y. Wada, T. Kastner, and M. J. Puma, 2017: Groundwater depletion embedded in
18 international food trade. *Nature*, **543**, 700.
- 19 Dallimer, M., L. C. Stringer, S. E. Orchard, P. Osano, G. Njoroge, C. Wen, and P. Gicheru, 2018:
20 Who uses sustainable land management practices and what are the costs and benefits? Insights
21 from Kenya. *L. Degrad. Dev.*, doi:10.1002/ldr.3001.
- 22 van Dam, J., M. Junginger, and A. P. C. Faaij, 2010: From the global efforts on certification of
23 bioenergy towards an integrated approach based on sustainable land use planning. *Renew.*
24 *Sustain. Energy Rev.*, **14**, 2445–2472, doi:10.1016/J.RSER.2010.07.010. [https://www-](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S1364032110001905)
25 [sciencedirect-com.ezp.sub.su.se/science/article/pii/S1364032110001905](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S1364032110001905) (Accessed May 23,
26 2018).
- 27 Damnyag, L., O. Saastamoinen, M. Appiah, and A. Pappinen, 2012: Role of tenure insecurity in
28 deforestation in Ghana's high forest zone. *For. Policy Econ.*, **14**, 90–98,
29 doi:10.1016/j.forpol.2011.08.006.
- 30 Dan, R., 2016: Optimal adaptation to extreme rainfalls in current and future climate. *Water Resour.*
31 *Res.*, **53**, 535–543, doi:10.1002/2016WR019718.
- 32 Daniel, S., 2011: Land Grabbing and Potential Implications for World Food Security. *Sustainable*
33 *Agricultural Development*, 25–42.
- 34 Darnhofer, I., 2014: Socio-technical transitions in farming: key concepts. *Transition pathways*
35 *towards sustainability in agriculture: case studies from Europe*.
- 36 Daron, J. D., and D. A. Stainforth, 2014: Assessing pricing assumptions for weather index insurance
37 in a changing climate. *Clim. Risk Manag.*, doi:10.1016/j.crm.2014.01.001.
- 38 Das, S., and J. R. Vincent, 2009: Mangroves protected villages and reduced death toll during Indian
39 super cyclone. *Proc. Natl. Acad. Sci.*, **106**, 7357–7360.
- 40 Dasgupta, P., A. P. Kinzig, and C. Perrings, 2013: The Value of Biodiversity. *Encyclopedia of*
41 *Biodiversity: Second Edition*.
- 42 Dasgupta, P., J. F. Morton, D. Dodman, B. Karapinar, F. Meza, M. G. Rivera-Ferre, A. Toure Sarr,
43 and K. E. Vincent, 2014: Rural Areas. *Climate Change 2014: Impacts, Adaptation, and*
44 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the*
45 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, 613–657.
- 46 Davies, J., 2017: *The Land in Drylands: Thriving in uncertainty through diversity*. Bonn, Germany,
47 18 pp. <https://knowledge.unccd.int/sites/default/files/2018-06/15>.

- 1 The%2BLand%2Bin%2BDrylands__J_Davies.pdf (Accessed April 15, 2019).
- 2 ———, C. Ogali, P. Laban, and G. Metternicht, 2015: *HOMING IN ON THE RANGE: Enabling*
3 *Investments for Sustainable Land Management*. www.iucn.org/drylands (Accessed November 3,
4 2018).
- 5 Davies, M., B. Guenther, J. Leavy, T. Mitchell, and T. Tanner, 2009: *Climate change adaptation,*
6 *disaster risk reduction, and social protection: complementary roles in agriculture and rural*
7 *growth?* 1-37 pp.
- 8 Davies, M., C. Béné, A. Arnall, T. Tanner, A. Newsham, and C. Coirolo, 2013: Promoting Resilient
9 Livelihoods through Adaptive Social Protection: Lessons from 124 programmes in South Asia.
10 *Dev. Policy Rev.*, doi:10.1111/j.1467-7679.2013.00600.x.
- 11 Davis, K. F., P. D’Odorico, and M. C. Rulli, 2014: Land grabbing: a preliminary quantification of
12 economic impacts on rurallivelihoods. *Popul. Environ.*, **36**, 180–192, doi:10.1007/s11111-014-
13 0215-2.
- 14 Davis, S. C., and Coauthors, 2013: Management swing potential for bioenergy crops. *GCB Bioenergy*,
15 doi:10.1111/gcbb.12042.
- 16 Debortoli, N. S., J. S. Sayles, D. G. Clark, and J. D. Ford, 2018: A systems network approach for
17 climate change vulnerability assessment. *Environ. Res. Lett.*, **13**, 104019, doi:10.1088/1748-
18 9326/aae24a. [http://stacks.iop.org/1748-
19 9326/13/i=10/a=104019?key=crossref.447573ce22ec275adaa5c597bc680a7e](http://stacks.iop.org/1748-9326/13/i=10/a=104019?key=crossref.447573ce22ec275adaa5c597bc680a7e) (Accessed
20 November 1, 2018).
- 21 DeClerck, F., 2016: IPBES: Biodiversity central to food security. *Nature*, **531**, 305.
- 22 Deininger, K., 2011: Challenges posed by the new wave of farmland investment. *J. Peasant Stud.*, **38**,
23 217–247, doi:10.1080/03066150.2011.559007.
24 [http://www.informaworld.com/smpp/title~content=t713673200URL:http://dx.doi.org/10.1080/0
25 3066150.2011.559007](http://www.informaworld.com/smpp/title~content=t713673200URL:http://dx.doi.org/10.1080/03066150.2011.559007)<http://www.informaworld.com/> (Accessed November 2, 2018).
- 26 Deininger, K., and O. Feder, 2009: Land registration, governance, and development: Evidence and
27 implications for policy. *World Bank Res. Obs.*, **24**, 233–266.
- 28 Deininger, K., and D. Byerlee, 2011: *The rise of large farms in land abundant countries: Do they*
29 *have a future?* 37 pp.
- 30 Dellasala, D. A., J. E. Williams, C. D. Williams, and J. F. Franklin, 2004: Beyond smoke and mirrors:
31 A synthesis of fire policy and science. *Conserv. Biol.*, **18**, 976–986, doi:10.1111/j.1523-
32 1739.2004.00529.x.
- 33 Dennison, P. E., S. C. Brewer, J. D. Arnold, and M. A. Moritz, 2014: Large wildfire trends in the
34 western United States, 1984-2011. *Geophys. Res. Lett.*, **41**, 2928–2933,
35 doi:10.1002/2014GL059576. <http://doi.wiley.com/10.1002/2014GL059576> (Accessed April 14,
36 2019).
- 37 Denton, F., and Coauthors, 2014: Climate-resilient pathways: Adaptation, mitigation, and sustainable
38 development. *Climate Change 2014 Impacts, Adaptation and Vulnerability: Part A: Global and*
39 *Sectoral Aspects*, 1101–1131.
- 40 Dercon, S., R. V. Hill, D. Clarke, I. Outes-Leon, and A. Seyoum Taffesse, 2014: Offering rainfall
41 insurance to informal insurance groups: Evidence from a field experiment in Ethiopia. *J. Dev.*
42 *Econ.*, **106**, 132–143, doi:10.1016/j.jdeveco.2013.09.006.
- 43 Deryng, D., D. Conway, N. Ramankutty, J. Price, and R. Warren, 2014: Global crop yield response to
44 extreme heat stress under multiple climate change futures. *Environ. Res. Lett.*,
45 doi:10.1088/1748-9326/9/3/034011.
- 46 Deryugina, T., 2013: Reducing the Cost of Ex Post Bailouts with Ex Ante Regulation: Evidence from
47 Building Codes. *SSRN Electron. J.*, doi:10.2139/ssrn.2314665.

- 1 <https://www.ssrn.com/abstract=2314665> (Accessed April 3, 2019).
- 2 Devereux, S., 2007: The impact of droughts and floods on food security and policy options to
3 alleviate negative effects. *Agricultural Economics*.
- 4 Dey Mayukh, 2018: Conserving river dolphins in a changing soundscape: acoustic and behavioural
5 responses of Ganges river dolphins to anthropogenic noise in the Ganges River, India. Tata
6 Institute of Fundamental Research, India, .
- 7 Dey, P., and A. Sakar, 2011: Revisiting indigenous farming knowledge of Jharkhand (India) for
8 conservation of natural resources and combating climate change. *Indian J Tradit Knowl*, **10**,
9 71–79. <http://nopr.niscair.res.in/handle/123456789/11067> (Accessed April 24, 2019).
- 10 Dharmadhikary, S., and Sandbhor, J., 2017: *National Inland Waterways in India: A Strategic Status*
11 *Report. Manthan, India.* 67 pp. [http://www.manthan-india.org/wp-](http://www.manthan-india.org/wp-content/uploads/2018/04/Strategic-Status-Report-on-Inland-Waterways-V5-26-Apr-17-FINAL.pdf)
12 [content/uploads/2018/04/Strategic-Status-Report-on-Inland-Waterways-V5-26-Apr-17-](http://www.manthan-india.org/wp-content/uploads/2018/04/Strategic-Status-Report-on-Inland-Waterways-V5-26-Apr-17-FINAL.pdf)
13 [FINAL.pdf](http://www.manthan-india.org/wp-content/uploads/2018/04/Strategic-Status-Report-on-Inland-Waterways-V5-26-Apr-17-FINAL.pdf).
- 14 DIAZ-CHAVEZ, R., 2015: Assessing sustainability for biomass energy production and use. 199–227,
15 doi:10.4324/9781315723273-15. [https://www.taylorfrancis-](https://www.taylorfrancis-com.ezp.sub.su.se/books/e/9781317527473/chapters/10.4324%2F9781315723273-15)
16 [com.ezp.sub.su.se/books/e/9781317527473/chapters/10.4324%2F9781315723273-15](https://www.taylorfrancis-com.ezp.sub.su.se/books/e/9781317527473/chapters/10.4324%2F9781315723273-15) (Accessed
17 October 15, 2018).
- 18 Diaz-Chavez, R. A., 2011: Assessing biofuels: Aiming for sustainable development or complying
19 with the market? *Energy Policy*, **39**, 5763–5769, doi:10.1016/J.ENPOL.2011.03.054.
20 <https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0301421511002436> (Accessed
21 May 23, 2018).
- 22 Díaz-Reviriego, I., Á. Fernández-Llamazares, P. L. Howard, J. L. Molina, and V. Reyes-García, 2017:
23 Fishing in the Amazonian Forest: A Gendered Social Network Puzzle. *Soc. Nat. Resour.*, **30**,
24 690–706, doi:10.1080/08941920.2016.1257079.
25 <https://www.tandfonline.com/doi/full/10.1080/08941920.2016.1257079> (Accessed April 9,
26 2019).
- 27 Diaz, R. J., and R. Rosenberg, 2008: Spreading dead zones and consequences for marine ecosystems.
28 *Science (80-.)*, **321**, 926–929.
- 29 Díaz, S., and Coauthors, 2015: The IPBES Conceptual Framework—connecting nature and people.
30 *Curr. Opin. Environ. Sustain.*, **14**, 1–16.
- 31 Diffenbaugh, N. S., T. W. Hertel, M. Scherer, and M. Verma, 2012: Response of corn markets to
32 climate volatility under alternative energy futures. *Nat. Clim. Chang.*,
33 doi:10.1038/nclimate1491.
- 34 Dillon, R. L., C. H. Tinsley, and W. J. Burns, 2014: Near-Misses and Future Disaster Preparedness.
35 *Risk Anal.*, **34**, 1907–1922, doi:10.1111/risa.12209.
36 <http://www.ncbi.nlm.nih.gov/pubmed/24773610> (Accessed April 3, 2019).
- 37 Distefano, T., F. Laio, L. Ridolfi, and S. Schiavo, 2018: Shock transmission in the international food
38 trade network. *PLoS One*, doi:10.1371/journal.pone.0200639.
- 39 Dittrich, R., A. Wreford, C. F. E. Topp, V. Eory, and D. Moran, 2017: A guide towards climate
40 change adaptation in the livestock sector: adaptation options and the role of robust decision-
41 making tools for their economic appraisal. *Reg. Environ. Chang.*, **17**, doi:10.1007/s10113-017-
42 1134-4.
- 43 Djalante, R., C. Holley, F. Thomalla, and M. Carnegie, 2013: Pathways for adaptive and integrated
44 disaster resilience. *Nat. Hazards*, doi:10.1007/s11069-013-0797-5.
- 45 Djoudi, H., and M. Brockhaus, 2011: Is adaptation to climate change gender neutral? Lessons from
46 communities dependent on livestock and forests in northern Mali. *Int. For. Rev.*, **13**, 123–135,
47 doi:10.1505/146554811797406606.

- 1 Djoudi, H., B. Locatelli, C. Vaast, K. Asher, M. Brockhaus, and B. Basnett Sijapati, 2016: Beyond
2 dichotomies: Gender and intersecting inequalities in climate change studies. *Ambio*, **45**, 248–
3 262, doi:10.1007/s13280-016-0825-2.
- 4 Domínguez, P., 2014: Current situation and future patrimonializing perspectives for the governance of
5 agro-pastoral resources in the Ait Ikis transhumants of the High Atlas (Morocco). 148–166,
6 doi:10.4324/9781315768014-14.
7 <https://www.taylorfrancis.com/books/e/9781317665175/chapters/10.4324/9781315768014-14>
8 (Accessed April 8, 2019).
- 9 Donato, D. C., J. B. Kauffman, D. Murdiyarso, S. Kurnianto, M. Stidham, and M. Kanninen, 2011:
10 Mangroves among the most carbon-rich forests in the tropics. *Nat. Geosci.*, **4**, 293.
- 11 Dooley, K., and S. Kartha, 2018: Land-based negative emissions: risks for climate mitigation and
12 impacts on sustainable development. *Int. Environ. Agreements Polit. Law Econ.*, **18**, 79–98,
13 doi:10.1007/s10784-017-9382-9. <http://link.springer.com/10.1007/s10784-017-9382-9>
14 (Accessed March 29, 2019).
- 15 Doss, C., C. Kovarik, A. Peterman, A. Quisumbing, and M. van den Bold, 2015a: Gender inequalities
16 in ownership and control of land in Africa: myth and reality. *Agric. Econ.*, **46**, 403–434,
17 doi:10.1111/agec.12171. <http://doi.wiley.com/10.1111/agec.12171> (Accessed April 2, 2019).
- 18 —, —, —, —, and —, 2015b: Gender inequalities in ownership and control of land in
19 Africa: myth and reality. *Agric. Econ.*, **46**, 403–434, doi:10.1111/agec.12171.
- 20 —, R. Meinzen-Dick, A. Quisumbing, and S. Theis, 2018a: Women in agriculture: Four myths.
21 *Glob. Food Sec.*, **16**, 69–74, doi:10.1016/J.GFS.2017.10.001.
22 <https://www.sciencedirect.com/science/article/pii/S2211912417300779> (Accessed April 2,
23 2019).
- 24 —, —, —, and —, 2018b: Women in agriculture: Four myths. *Glob. Food Sec.*, **16**, 69–74,
25 doi:10.1016/J.GFS.2017.10.001.
- 26 Dougill, A. J., E. D. G. Fraser, and M. S. Reed, 2011: Anticipating Vulnerability to Climate Change
27 in Dryland Pastoral Systems : Using Dynamic Systems Models for the Kalahari. *Ecol. Soc.*, **15**,
28 doi:10.2307/26268132. <http://www.jstor.org/stable/26268132> (Accessed May 22, 2018).
- 29 Douglas, I., K. Alam, M. Maghenda, Y. Mcdonnell, L. Mclean, and J. Campbell, 2008: Unjust waters:
30 climate change, flooding and the urban poor in Africa. *Environ. Urban.*, **20**, 187–205,
31 doi:10.1177/0956247808089156. <http://journals.sagepub.com/doi/10.1177/0956247808089156>
32 (Accessed May 17, 2018).
- 33 Dow, K., F. Berkhout, and B. L. Preston, 2013: Limits to adaptation to climate change: A risk
34 approach. *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2013.07.005.
- 35 Dowdy, A. J., and A. Pepler, 2018: Pyroconvection Risk in Australia: Climatological Changes in
36 Atmospheric Stability and Surface Fire Weather Conditions. *Geophys. Res. Lett.*, **45**, 2005–
37 2013, doi:10.1002/2017GL076654. <http://doi.wiley.com/10.1002/2017GL076654> (Accessed
38 April 14, 2019).
- 39 Downing, T. ., 2012: Views of the frontiers in climate change adaptation economics. *Wiley*
40 *Interdiscip. Rev. Clim. Chang.*, **3**, 161–170, doi:10.1002/wcc.157.
- 41 Driscoll, D. A., M. Bode, R. A. Bradstock, D. A. Keith, T. D. Penman, and O. F. Price, 2016:
42 Resolving future fire management conflicts using multicriteria decision making. *Conserv. Biol.*,
43 **30**, 196–205, doi:10.1111/cobi.12580.
- 44 Dryzek, J. S., and J. Pickering, 2017: Deliberation as a catalyst for reflexive environmental
45 governance. *Ecol. Econ.*, **131**, 353–360, doi:10.1016/j.ecolecon.2016.09.011.
- 46 Duchelle, A. E., and Coauthors, 2017: Balancing carrots and sticks in REDD+: implications for social
47 safeguards. *Ecol. Soc.*, **22**, art2, doi:10.5751/ES-09334-220302.
48 <https://www.ecologyandsociety.org/vol22/iss3/art2/> (Accessed October 15, 2018).

- 1 Duguma, L. A., P. A. Minang, O. E. Freeman, and H. Hager, 2014a: System wide impacts of fuel
2 usage patterns in the Ethiopian highlands: Potentials for breaking the negative reinforcing
3 feedback cycles. doi:10.1016/j.esd.2014.03.004. <http://dx.doi.org/10.1016/j.esd.2014.03.004>
4 (Accessed April 14, 2019).
- 5 —, —, —, and —, 2014b: System wide impacts of fuel usage patterns in the Ethiopian
6 highlands: Potentials for breaking the negative reinforcing feedback cycles.
7 doi:10.1016/j.esd.2014.03.004.
- 8 Duguma, L. A., P. A. Minang, D. Foundjem-Tita, P. Makui, and S. M. Piabuo, 2018: Prioritizing
9 enablers for effective community forestry in Cameroon. *Ecol. Soc.*, doi:10.5751/ES-10242-
10 230301.
- 11 Durigan, G., and J. A. Ratter, 2016: The need for a consistent fire policy for Cerrado conservation. *J.*
12 *Appl. Ecol.*, **53**, 11–15, doi:10.1111/1365-2664.12559.
- 13 Duru, M., O. Therond, and M. Fares, 2015: Designing agroecological transitions; A review. *Agron.*
14 *Sustain. Dev.*, doi:10.1007/s13593-015-0318-x.
- 15 Dwyer, J., and I. Hodge, 2016: Governance structures for social-ecological systems: Assessing
16 institutional options against a social residual claimant. *Environ. Sci. Policy*,
17 doi:10.1016/j.envsci.2016.07.017.
- 18 Eakin, H. C., Cognitive and institutional influences on farmers' adaptive capacity: insights into
19 barriers and opportunities for transformative change in central Arizona.
- 20 Easdale, M. H., 2016: Zero net livelihood degradation-the quest for a multidimensional protocol to
21 combat desertification. *SOIL*, **2**, 129–134, doi:10.5194/soil-2-129-2016.
- 22 Easterly, W., 2008a: Institutions: Top Down or Bottom Up? *Am. Econ. Rev.*, **98**, 95–99,
23 doi:10.1257/aer.98.2.95. <http://pubs.aeaweb.org/doi/10.1257/aer.98.2.95> (Accessed May 22,
24 2018).
- 25 —, 2008b: Institutions: Top down or bottom up? *Am. Econ. Rev.*, **98**, 95–99,
26 doi:10.1257/aer.98.2.95. <http://pubs.aeaweb.org/doi/10.1257/aer.98.2.95> (Accessed May 23,
27 2018).
- 28 EEA, 2016: *Urban adaptation to climate change in Europe : Transforming Cities in a changing*
29 *climate*. Copenhagen, Denmark, 135 pp. [https://www.eea.europa.eu/publications/urban-](https://www.eea.europa.eu/publications/urban-adaptation-2016)
30 [adaptation-2016](https://www.eea.europa.eu/publications/urban-adaptation-2016).
- 31 Ehara, M., K. Hyakumura, R. Sato, K. Kurosawa, K. Araya, H. Sokh, and R. Kohsaka, 2018:
32 Addressing Maladaptive Coping Strategies of Local Communities to Changes in Ecosystem
33 Service Provisions Using the DPSIR Framework. *Ecol. Econ.*,
34 doi:10.1016/j.ecolecon.2018.03.008.
- 35 El-Naggar, A., and Coauthors, 2018: Influence of soil properties and feedstocks on biochar potential
36 for carbon mineralization and improvement of infertile soils. *Geoderma*, **332**, 100–108,
37 doi:10.1016/j.geoderma.2018.06.017. [https://www.sciencedirect-](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0016706118300715)
38 [com.ezp.sub.su.se/science/article/pii/S0016706118300715](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0016706118300715) (Accessed April 14, 2019).
- 39 ELD Initiative, 2015: *The value of land: Prosperous lands and positive rewards through sustainable*
40 *land management*. Bonn,.
- 41 Eldridge, D. J., M. A. Bowker, F. T. Maestre, E. Roger, J. F. Reynolds, and W. G. Whitford, 2011:
42 Impacts of shrub encroachment on ecosystem structure and functioning: towards a global
43 synthesis. *Ecol. Lett.*, **14**, 709–722, doi:10.1111/j.1461-0248.2011.01630.x.
44 <http://www.ncbi.nlm.nih.gov/pubmed/21592276> (Accessed May 18, 2018).
- 45 Eling, M., S. Pradhan, and J. T. Schmit, 2014: The Determinants of Microinsurance Demand. *Geneva*
46 *Pap. Risk Insur. - Issues Pract.*, **39**, 224–263, doi:10.1057/gpp.2014.5.
47 <http://link.springer.com/10.1057/gpp.2014.5> (Accessed April 3, 2019).

- 1 Ellis, F., S. Devereux, and P. White, 2009: Social Protection in Africa. *Enterp. Dev. Microfinance*, **20**,
2 158–160, doi:10.3362/1755-1986.2009.015.
- 3 Ellison, D., and Coauthors, 2017: Trees, forests and water: Cool insights for a hot world. *Glob.*
4 *Environ. Chang.*, **43**, 51–61.
- 5 Elmqvist, T., C. Folke, M. Nyström, G. Peterson, J. Bengtsson, B. Walker, and J. Norberg, 2003:
6 Response diversity, ecosystem change, and resilience. *Front. Ecol. Environ.*, **1**, 488–494,
7 doi:10.1890/1540-9295(2003)001[0488:rdecar]2.0.co;2.
- 8 ———, M. Tuvendal, J. Krishnaswamy, and K. Hylander, 2013: Managing trade-offs in ecosystem
9 services. *Values, payments institutions Ecosyst. Manag. Cheltenham Edward Elgar Publ.*, 70–
10 89.
- 11 Emerson, K., and A. K. Gerlak, 2014: Adaptation in Collaborative Governance Regimes. *Environ.*
12 *Manage.*, doi:10.1007/s00267-014-0334-7.
- 13 Endres, J., R. Diaz-Chavez, S. R. Kaffka, L. Pelkmans, J. E. A. Seabra, and A. Walter, 2015:
14 Sustainability Certification. *Bioenergy & Sustainability: Bridging the gaps*, L. Souza, G. M.,
15 Victoria, R., Joly, C., Verdade, Ed., SCOPE, Paris
16 http://bioenfapesp.org/scopebioenergy/images/chapters/bioen-scope_chapter19.pdf (Accessed
17 May 23, 2018).
- 18 Enfors, E. I., and L. J. Gordon, 2008: Dealing with drought: The challenge of using water system
19 technologies to break dryland poverty traps. *Glob. Environ. Chang.*, **18**, 607–616,
20 doi:10.1016/J.GLOENVCHA.2008.07.006.
21 <https://www.sciencedirect.com/science/article/abs/pii/S0959378008000575> (Accessed April 8,
22 2019).
- 23 Engel, S., and A. Muller, 2016: Payments for environmental services to promote “climate-smart
24 agriculture”? Potential and challenges. *Agric. Econ.*, **47**, 173–184, doi:10.1111/agec.12307.
25 <http://doi.wiley.com/10.1111/agec.12307> (Accessed December 28, 2017).
- 26 Englund, O., and G. Berndes, 2015: How do sustainability standards consider biodiversity? *Wiley*
27 *Interdiscip. Rev. Energy Environ.*, **4**, 26–50, doi:10.1002/wene.118.
28 <http://doi.wiley.com/10.1002/wene.118> (Accessed April 3, 2019).
- 29 Ens, E. J., and Coauthors, 2015: Indigenous biocultural knowledge in ecosystem science and
30 management: Review and insight from Australia. *Biol. Conserv.*, **181**, 133–149,
31 doi:10.1016/J.BIOCON.2014.11.008.
32 <https://www.sciencedirect.com/science/article/abs/pii/S0006320714004339> (Accessed April 6,
33 2019).
- 34 Ensor, J., 2011: *Uncertain futures : adapting development to a changing climate*. Practical Action
35 Pub, 108 pp.
- 36 ———, and B. Harvey, 2015: Social learning and climate change adaptation: evidence for international
37 development practice. *Wiley Interdiscip. Rev. Clim. Chang.*, **6**, 509–522, doi:10.1002/wcc.348.
38 <http://doi.wiley.com/10.1002/wcc.348> (Accessed January 2, 2018).
- 39 Eory, V., C. F. E. Topp, A. Butler, and D. Moran, 2018: Addressing Uncertainty in Efficient
40 Mitigation of Agricultural Greenhouse Gas Emissions. *J. Agric. Econ.*, **69**, 627–645,
41 doi:10.1111/1477-9552.12269.
- 42 EPA, 2018: *Ireland’s Final Greenhouse Gas Emissions: 1990 - 2016*. Dublin,
43 [http://www.epa.ie/pubs/reports/air/airemissions/ghgemissions2016/Report_GHG_1990-2016](http://www.epa.ie/pubs/reports/air/airemissions/ghgemissions2016/Report_GHG_1990-2016_April_for_Website-v3.pdf)
44 [April_for Website-v3.pdf](http://www.epa.ie/pubs/reports/air/airemissions/ghgemissions2016/Report_GHG_1990-2016_April_for_Website-v3.pdf).
- 45 Epanchin-Niell, R. S., M. B. Hufford, C. E. Asian, J. P. Sexton, J. D. Port, and T. M. Waring, 2010:
46 Controlling invasive species in complex social landscapes. *Front. Ecol. Environ.*,
47 doi:10.1890/090029.
- 48 Era Consultancy, 2006: *The Environment Impact Assessment (EIA) Notification, 2006*. New Delhi,

- 1 http://www.eraconsultancy.org.in/pdf/EIA_Notification_2006.pdf (Accessed April 14, 2019).
- 2 Eriksen, S., and Coauthors, 2011: When not every response to climate change is a good one:
3 Identifying principles for sustainable adaptation. *Clim. Dev.*, **3**, 7–20,
4 doi:10.3763/cdev.2010.0060.
- 5 Erkossa, T., A. Wudneh, B. Desalegn, and G. Taye, 2015: Linking soil erosion to on-site financial
6 cost: Lessons from watersheds in the Blue Nile basin. *Solid Earth*, **6**, 765–774, doi:10.5194/se-
7 6-765-2015.
- 8 Estrin, D., 2016: *LIMITING DANGEROUS CLIMATE CHANGE THE CRITICAL ROLE OF*
9 *CITIZEN SUITS AND DOMESTIC COURTS — DESPITE THE PARIS AGREEMENT*. 36 pp.
- 10 —, and S. V. Tan, 2016: Thinking Outside the Boat about Climate Change Loss and Damage.
11 [https://www.cigionline.org/publications/thinking-outside-boat-about-climate-change-loss-and-](https://www.cigionline.org/publications/thinking-outside-boat-about-climate-change-loss-and-damage)
12 [damage](https://www.cigionline.org/publications/thinking-outside-boat-about-climate-change-loss-and-damage) (Accessed May 25, 2018).
- 13 Estrin, S., and M. Prevezer, 2011: The role of informal institutions in corporate governance: Brazil,
14 Russia, India, and China compared. *Asia Pacific J. Manag.*, **28**, 41–67, doi:10.1007/s10490-010-
15 9229-1.
- 16 Etkin, D., J. Medalye, and K. Higuchi, 2012: Climate warming and natural disaster management: An
17 exploration of the issues. *Clim. Change*, doi:10.1007/s10584-011-0259-6.
- 18 European Commission, 2012: *Renewable energy progress and biofuels sustainability*.
19 http://ec.europa.eu/energy/renewables/reports/doc/2013_renewable_energy_progress.pdf
20 (Accessed October 15, 2018).
- 21 European Union, 2018: *DIRECTIVES DIRECTIVE (EU) 2018/2001 OF THE EUROPEAN*
22 *PARLIAMENT AND OF THE COUNCIL of 11 December 2018 on the promotion of the use of*
23 *energy from renewable sources (recast) (Text with EEA relevance)*. [https://eur-](https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32018L2001&from=EN)
24 [lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32018L2001&from=EN](https://eur-lex.europa.eu/legal-content/EN/TXT/PDF/?uri=CELEX:32018L2001&from=EN) (Accessed
25 April 8, 2019).
- 26 Ewel, K., R. TWILLEY, and J. I. N. Ong, 1998: Different kinds of mangrove forests provide different
27 goods and services. *Glob. Ecol. Biogeogr. Lett.*, **7**, 83–94.
- 28 Faaij, A. P., 2018: *Securing sustainable resource availability of biomass for energy applications in*
29 *Europe; review of recent literature. The role of biomass for energy and materials for GHG*
30 *mitigation from a global and European perspective*. [https://bioenergyeurope.org/wp-](https://bioenergyeurope.org/wp-content/uploads/2018/11/Bioenergy-Europe-EU-Biomass-Resources-André-Faaij-Final.pdf)
31 [content/uploads/2018/11/Bioenergy-Europe-EU-Biomass-Resources-André-Faaij-Final.pdf](https://bioenergyeurope.org/wp-content/uploads/2018/11/Bioenergy-Europe-EU-Biomass-Resources-André-Faaij-Final.pdf)
32 (Accessed April 8, 2019).
- 33 Fadaïro, O. S., R. Calland, Y. Mulugetta, and J. Olawoye, 2017: A Corruption Risk Assessment for
34 Reducing Emissions from Deforestation and Forest Degradation in Nigeria. *Int. J. Clim. Chang.*
35 *Impacts Responses*, **10**, 1–21, doi:10.18848/1835-7156/CGP/v10i01/1-21.
36 [https://cgscholar.com/bookstore/works/a-corruption-risk-assessment-for-reducing-emissions-](https://cgscholar.com/bookstore/works/a-corruption-risk-assessment-for-reducing-emissions-from-deforestation-and-forest-degradation-in-nigeria)
37 [from-deforestation-and-forest-degradation-in-nigeria](https://cgscholar.com/bookstore/works/a-corruption-risk-assessment-for-reducing-emissions-from-deforestation-and-forest-degradation-in-nigeria) (Accessed May 23, 2018).
- 38 Fairbairn, M., 2015: Foreignization, Financialization and Land Grab Regulation. *J. Agrar. Chang.*, **15**,
39 581–591, doi:10.1111/joac.12112.
- 40 Falkenmark, M., 2001: The greatest water problem: the inability to link environmental security, water
41 security and food security. *Int. J. Water Resour. Dev.*, **17**, 539–554.
- 42 Falkowski, T. B., S. A. W. Diemont, A. Chankin, and D. Douterlungne, 2016: Lacandon Maya
43 traditional ecological knowledge and rainforest restoration: Soil fertility beneath six agroforestry
44 system trees. *Ecol. Eng.*, **92**, 210–217, doi:10.1016/J.ECOLENG.2016.03.002.
45 <https://www.sciencedirect.com/science/article/abs/pii/S0925857416301513> (Accessed April 6,
46 2019).
- 47 Fameree, C., 2016: Political contestations around land deals: insights from Peru. *Can. J. Dev. Stud.*
48 *Can. D ETUDES DU DEVELOPPEMENT*, **37**, 541–559, doi:10.1080/02255189.2016.1175340.

- 1 Fankhauser, S., 2017: Adaptation to Climate Change.
- 2 FAO, 2010: *Climate-Smart Agriculture: Policies, Practices and Financing for Food Security,*
3 *Adaptation and Mitigation.* Rome,.
- 4 —, 2011a: *State of Food and Agriculture 2010-2011.* 147 pp.
- 5 —, 2011b: *THE STATE OF FOOD AND AGRICULTURE WOMEN IN AGRICULTURE 2010-11*
6 *.Closing the gender gap for development.* <http://www.fao.org/catalog/inter-e.htm> (Accessed
7 April 2, 2019).
- 8 —, 2011c: *THE STATE OF FOOD AND AGRICULTURE WOMEN IN AGRICULTURE 2010-11*
9 *.Closing the gender gap for development.*
- 10 —, 2015a: Voluntary Guidelines for Securing Sustainable Small-Scale Fisheries in the Context of
11 Food Security and Poverty Eradication. *Food Agric. Organ. United Nations.*,
12 <http://www.fao.org/3/a-i4356en.pdf> (Accessed November 1, 2018).
- 13 —, 2015b: *The impact of disasters on agriculture and food security.* ITALY,
14 <http://www.fao.org/3/a-i5128e.pdf> (Accessed October 3, 2018).
- 15 —, 2015c: *Gender and Land Statistics Recent developments in FAO's Gender and Land Rights*
16 *Database.* ROME, www.fao.org/publications (Accessed April 4, 2019).
- 17 —, 2015d: *Gender and Land Statistics Recent developments in FAO's Gender and Land Rights*
18 *Database.* ROME,.
- 19 —, 2016: *The agriculture sectors in the Intended Nationally Determined Contributions: Analysis.*
20 Rome, Italy, <http://www.fao.org/3/a-i5687e.pdf>.
- 21 —, 2017a: *The future of food and agriculture: Trends and challenges.* 180 pp.
22 <http://www.fao.org/3/a-i6583e.pdf>.
- 23 —, 2017b: FAO Cereal Supply and Demand Brief | FAO | Food and Agriculture Organization of
24 the United Nations. *Fao*, 1.
- 25 Farber, D. A., 2015: Coping with uncertainty: Cost-benefit analysis, the precautionary principle, and
26 climate change. *Washingt. Law Rev.*, doi:10.1525/sp.2007.54.1.23.
- 27 Farfan, J., and C. Breyer, 2017: Structural changes of global power generation capacity towards
28 sustainability and the risk of stranded investments supported by a sustainability indicator. *J.*
29 *Clean. Prod.*, **141**, 370–384.
- 30 Fawcett, A. ., L. . Clarke, S. Rausch, and J. P. Weyant, 2014: Overview of EMF 24 Policy Scenarios.
31 *Energy J.*, **35**, 33–60.
- 32 Faye, B., and Coauthors, 2018: Impacts of 1.5 versus 2.0 °c on cereal yields in the West African
33 Sudan Savanna. *Environ. Res. Lett.*, doi:10.1088/1748-9326/aaab40.
- 34 Fearnside, P. M., 2015: Deforestation soars in the Amazon. *Nature*, **521**, 423–423,
35 doi:10.1038/521423b. <http://www.nature.com/articles/521423b> (Accessed April 23, 2019).
- 36 Feliciano, D., C. Hunter, B. Slee, and P. Smith, 2014: ScienceDirect Climate change mitigation
37 options in the rural land use sector : Stakeholders ' perspectives on barriers , enablers and
38 the role of policy in North East Scotland. *Environ. Sci. Policy*, **44**, 26–38,
39 doi:10.1016/j.envsci.2014.07.010. <http://dx.doi.org/10.1016/j.envsci.2014.07.010>.
- 40 Fellmann, T., P. Witzke, F. Weiss, B. Van Doorslaer, D. Drabik, I. Huck, G. Salputra, and T. Jansson,
41 2018: Major challenges of integrating agriculture into climate change mitigation policy
42 frameworks. 451–468, doi:10.1007/s11027-017-9743-2.
- 43 Fencl, J. S., M. E. Mather, K. H. Costigan, and M. D. Daniels, 2015: How big of an effect do small
44 dams have? Using geomorphological footprints to quantify spatial impact of low-head dams and
45 identify patterns of across-dam variation. *PLoS One*, **10**, e0141210.

- 1 Ferguson, G., and T. Gleeson, 2012: Vulnerability of coastal aquifers to groundwater use and climate
2 change. *Nat. Clim. Chang.*, **2**, 342–345, doi:10.1038/nclimate1413.
3 <http://www.nature.com/articles/nclimate1413> (Accessed March 29, 2019).
- 4 Fernandes, K., L. Verchot, W. Baethgen, V. Gutierrez-Velez, M. Pinedo-Vasquez, and C. Martius,
5 2017: Heightened fire probability in Indonesia in non-drought conditions: the effect of
6 increasing temperatures. *Environ. Res. Lett.*, **12**, 054002, doi:10.1088/1748-9326/aa6884.
7 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=5/a=054002?key=crossref.cb943b6a65dbd8384efb770f0ca7d8ed)
8 [9326/12/i=5/a=054002?key=crossref.cb943b6a65dbd8384efb770f0ca7d8ed](http://stacks.iop.org/1748-9326/12/i=5/a=054002?key=crossref.cb943b6a65dbd8384efb770f0ca7d8ed) (Accessed April 14,
9 2019).
- 10 Fernandez-Gimenez, M. E., 2000: THE ROLE OF MONGOLIAN NOMADIC PASTORALISTS'
11 ECOLOGICAL KNOWLEDGE IN RANGELAND MANAGEMENT. *Ecol. Appl.*, **10**, 1318–
12 1326, doi:10.1890/1051-0761(2000)010[1318:TROMNP]2.0.CO;2.
13 [https://esajournals.onlinelibrary.wiley.com/doi/abs/10.1890/1051-](https://esajournals.onlinelibrary.wiley.com/doi/abs/10.1890/1051-0761(2000)010%5B1318:TROMNP%5D2.0.CO%3B2)
14 [0761\(2000\)010%5B1318:TROMNP%5D2.0.CO%3B2](https://esajournals.onlinelibrary.wiley.com/doi/abs/10.1890/1051-0761(2000)010%5B1318:TROMNP%5D2.0.CO%3B2) (Accessed April 8, 2019).
- 15 Few, R., and M. G. L. Tebboth, 2018: Recognising the dynamics that surround drought impacts. *J.*
16 *Arid Environ.*,.
- 17 Feyen, E., R. Lester, and R. Rocha, 2011: What drives the development of the insurance sector ? an
18 empirical analysis based on a panel of developed and developing countries. *Policy Res. Work.*
19 *Pap. Ser.*, <https://ideas.repec.org/p/wbk/wbrwps/5572.html> (Accessed April 3, 2019).
- 20 Filatova, T., 2014: Market-based instruments for flood risk management: A review of theory, practice
21 and perspectives for climate adaptation policy. *Environ. Sci. Policy*, **37**, 227–242,
22 doi:10.1016/j.envsci.2013.09.005.
- 23 Filiberto, B. D., E. Wethington, and K. Pillemer, 2010: Older People and Climate Change :
24 Vulnerability and Health Effects. *Generations*, **33**, 19–26.
- 25 Findell, K. L., A. Berg, P. Gentine, J. P. Krasting, B. R. Lintner, S. Malyshev, J. A. Santanello, and E.
26 Shevliakova, 2017: The impact of anthropogenic land use and land cover change on regional
27 climate extremes. *Nat. Commun.*, **8**, 989, doi:10.1038/s41467-017-01038-w.
- 28 Finley-Brook, M., 2007: Indigenous land tenure insecurity fosters illegal logging in Nicaragua. *Int.*
29 *For. Rev.*, **9**, 850–864, doi:10.1505/ifor.9.4.850.
- 30 Fischer, C., and R. G. Newell, 2008: Environmental and technology policies for climate mitigation. *J.*
31 *Environ. Econ. Manage.*, **55**, 142–162, doi:10.1016/j.jeem.2007.11.001.
- 32 Fischer, E. M., and R. Knutti, 2015: Anthropogenic contribution to global occurrence of heavy-
33 precipitation and high-temperature extremes. *Nat. Clim. Chang.*, **5**, 560–564,
34 doi:10.1038/nclimate2617. <http://www.nature.com/articles/nclimate2617> (Accessed April 14,
35 2019).
- 36 Fischer, J., and Coauthors, 2017: Reframing the Food–Biodiversity Challenge. *Trends Ecol. Evol.*,
37 doi:10.1016/j.tree.2017.02.009.
- 38 Fishman, R., N. Devineni, and S. Raman, 2015: Can improved agricultural water use efficiency save
39 India's groundwater? *Environ. Res. Lett.*, doi:10.1088/1748-9326/10/8/084022.
- 40 Fiszbein, A., R. Kanbur, and R. Yemtsov, 2014: Social protection and poverty reduction: Global
41 patterns and some targets. *World Dev.*, **61**, 167–177, doi:10.1016/j.worlddev.2014.04.010.
- 42 Flahaux, M.-L., and H. De Haas, 2016: African migration: trends, patterns, drivers. *Comp. Migr.*
43 *Stud.*, doi:10.1186/s40878-015-0015-6.
- 44 Fleskens, Luuk, Stringer, L. ., 2014: Land management and policy responses to mitigate
45 desertification and land degradation. *L. Degrad. Dev.*, **25**, 1–4, doi:10.1002/ldr.2272.
46 <http://doi.wiley.com/10.1002/ldr.2272> (Accessed May 23, 2018).
- 47 Fleskens, L., D. Nainggolan, and L. C. Stringer, 2014: An Exploration of Scenarios to Support

- 1 Sustainable Land Management Using Integrated Environmental Socio-economic Models.
2 *Environ. Manage.*, **54**, 1005–1021, doi:10.1007/s00267-013-0202-x.
- 3 Fletcher, A. J., 2017: Maybe tomorrow will be better?: Gender and farm work in a changing climate.
4 *Climate change and gender in rich countries: Work, public policy and action*, M. Cohen, Ed.,
5 Routledge., Abingdon, Oxon; New York, NY., 185–198.
- 6 Fletcher, A. J., 2018: What works for women in agriculture? *Women in Agriculture Worldwide: Key*
7 *Issues and Practical Approaches*, A.J. Fletcher and W. Kubik, Eds., Routledge, London and
8 New York, 257–268 [https://www.routledge.com/Women-in-Agriculture-Worldwide-Key-issues-](https://www.routledge.com/Women-in-Agriculture-Worldwide-Key-issues-and-practical-approaches-1st/Fletcher-Kubik/p/book/9781472473080)
9 [and-practical-approaches-1st/Fletcher-Kubik/p/book/9781472473080](https://www.routledge.com/Women-in-Agriculture-Worldwide-Key-issues-and-practical-approaches-1st/Fletcher-Kubik/p/book/9781472473080) (Accessed April 2, 2019).
- 10 Fletcher, A. J., and E. Knuttila, 2016: Gendering change: Canadian farm women respond to drought.
11 *Vulnerability and adaptation to drought: The Canadian prairies and South America*, H. Diaz,
12 M. Hurlbert, and J. Warren, Eds., University of Calgary Press, Calgary, AB, 159–177.
- 13 Fleurbaey, M., and Coauthors, 2014: Sustainable Development and Equity. *Climate Change 2014:*
14 *Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment*
15 *Report of the Intergovernmental Panel on Climate Change*, 283–350.
- 16 Floater, G., P. Rode, B. Friedel, and A. Robert, 2014: Steering Urban Growth: Governance, Policy
17 and Finance. *LSE Cities*.
- 18 Folke, C., S. R. Carpenter, B. Walker, M. Scheffer, T. Chapin, and J. Rockström, 2010: Resilience
19 thinking: Integrating resilience, adaptability and transformability. *Ecol. Soc.*, **15**,
20 doi:10.5751/ES-03610-150420.
- 21 Fontaine, J. J., 2011: Improving our legacy: Incorporation of adaptive management into state wildlife
22 action plans. *J. Environ. Manage.*, **92**, 1403–1408, doi:10.1016/j.jenvman.2010.10.015.
- 23 Ford, A., R. Dawson, P. Blythe, and S. Barr, 2018: Land-use transport models for climate change
24 mitigation and adaptation planning. *J. Transp. Land Use*, **11**, 83–101,
25 doi:10.5198/jtlu.2018.1209. <http://dx.doi.org/10.5198/jtlu.2018.1209> (Accessed February 25,
26 2019).
- 27 Ford, J. D., and T. Pearce, 2010: What we know, do not know, and need to know about climate
28 change vulnerability in the western Canadian Arctic: A systematic literature review. *Environ.*
29 *Res. Lett.*, doi:10.1088/1748-9326/5/1/014008.
- 30 —, and C. Goldhar, 2012: Climate change vulnerability and adaptation in resource dependent
31 communities: A case study from West Greenland. *Clim. Res.*, **54**, 181–196,
32 doi:10.3354/cr01118.
- 33 Ford, J. D., and L. Berrang-Ford, 2016: The 4Cs of adaptation tracking: consistency, comparability,
34 comprehensiveness, coherency. *Mitig. Adapt. Strateg. Glob. Chang.*, **21**, 839–859,
35 doi:10.1007/s11027-014-9627-7. <https://doi.org/10.1007/s11027-014-9627-7>.
- 36 Ford, J. D., L. Cameron, J. Rubis, M. Maillet, D. Nakashima, A. C. Willox, and T. Pearce, 2016:
37 Including indigenous knowledge and experience in IPCC assessment reports. *Nat. Clim. Chang.*,
38 **6**, 349–353, doi:10.1038/nclimate2954. <http://www.nature.com/articles/nclimate2954> (Accessed
39 April 9, 2019).
- 40 Forsyth, T., 2018: Is resilience to climate change socially inclusive? Investigating theories of change
41 processes in Myanmar. *World Dev.*, **111**, 13–26, doi:10.1016/j.worlddev.2018.06.023.
42 <https://linkinghub.elsevier.com/retrieve/pii/S0305750X18302158> (Accessed April 16, 2019).
- 43 Foudi, S., and K. Erdlenbruch, 2012: The role of irrigation in farmers’ risk management strategies in
44 France. *Eur. Rev. Agric. Econ.*, **39**, 439–457.
- 45 Foxon, T. J., M. S. Reed, and L. C. Stringer, 2009: Governing Long-Term Social – Ecological
46 Change : What Can the Adaptive Management and Transition Management Approaches Learn
47 from Each Other ? *Change*, **20**, 3–20, doi:10.1002/eet.

- 1 Fra.Paleo, U., 2015: *Risk governance: The articulation of Hazard, politics and ecology*.
- 2 Frank, S., and Coauthors, 2017: Reducing greenhouse gas emissions in agriculture without
3 compromising food security? *Environ. Res. Lett.*, **12**, 105004, doi:10.1088/1748-9326/aa8c83.
4 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=10/a=105004?key=crossref.000b6ab1748d1b0af07f66e0f496f0a3)
5 [9326/12/i=10/a=105004?key=crossref.000b6ab1748d1b0af07f66e0f496f0a3](http://stacks.iop.org/1748-9326/12/i=10/a=105004?key=crossref.000b6ab1748d1b0af07f66e0f496f0a3) (Accessed
6 December 13, 2017).
- 7 Franz, M., N. Schlitz, and K. P. Schumacher, 2017: Globalization and the water-energy-food nexus -
8 Using the global production networks approach to analyze society-environment relations.
9 *Environmental Science and Policy*.
- 10 Fraser, E. D. G., W. Mabee, and F. Figge, 2005: A framework for assessing the vulnerability of food
11 systems to future shocks. *Futures*, **37**, 465–479, doi:10.1016/J.FUTURES.2004.10.011.
12 <https://www.sciencedirect.com/science/article/pii/S001632870400151X> (Accessed May 17,
13 2018).
- 14 Fraser, E. D. G., A. J. Dougill, K. Hubacek, C. H. Quinn, J. Sendzimir, and M. Termansen, 2011:
15 Assessing Vulnerability to Climate Change in Dryland Livelihood Systems: Conceptual
16 Challenges and Interdisciplinary Solutions. *Ecol. Soc.*, **16**, art3, doi:10.5751/ES-03402-160303.
17 <http://www.ecologyandsociety.org/vol16/iss3/art3/> (Accessed May 22, 2018).
- 18 Fre, Z., 2018: *Knowledge Sovereignty Among African Cattle Herders*. UCL Press.,
19 [https://www.ucl.ac.uk/bartlett/development/publications/2018/jun/knowledge-sovereignty-](https://www.ucl.ac.uk/bartlett/development/publications/2018/jun/knowledge-sovereignty-among-african-cattle-herders)
20 [among-african-cattle-herders](https://www.ucl.ac.uk/bartlett/development/publications/2018/jun/knowledge-sovereignty-among-african-cattle-herders) (Accessed April 24, 2019).
- 21 Frechette, A., C. Ginsburg, W. Walker, S. Gorelik, S. Keene, C. Meyer, K. Reytar, and P. Veit, 2018:
22 *A Global Baseline of Carbon Storage in Collective Lands*. Washington D C,
23 [https://rightsandresources.org/wp-content/uploads/2018/09/A-Global-Baseline_RRI_Sept-](https://rightsandresources.org/wp-content/uploads/2018/09/A-Global-Baseline_RRI_Sept-2018.pdf)
24 [2018.pdf](https://rightsandresources.org/wp-content/uploads/2018/09/A-Global-Baseline_RRI_Sept-2018.pdf) (Accessed April 13, 2019).
- 25 Fredriksson, P. G., and E. Neumayer, 2016: Corruption and Climate Change Policies: Do the Bad Old
26 Days Matter? *Environ. Resour. Econ.*, **63**, 451–469, doi:10.1007/s10640-014-9869-6.
27 <http://link.springer.com/10.1007/s10640-014-9869-6> (Accessed October 5, 2018).
- 28 Freebairn, J., 2016: A Comparison of Policy Instruments to Reduce Greenhouse Gas Emissions *.
29 *Econ. Pap.*, **35**, 204–215, doi:10.1111/1759-3441.12141.
- 30 French, S., 2013: Cynefin, statistics and decision analysis. *J. Oper. Res. Soc.*,
31 doi:10.1057/jors.2012.23.
- 32 French, S., 2015: Cynefin: Uncertainty, small worlds and scenarios. *J. Oper. Res. Soc.*, **66**, 1635–
33 1645, doi:10.1057/jors.2015.21.
- 34 Fridahl, M., and B. O. Linnér, 2016: Perspectives on the Green Climate Fund: possible compromises
35 on capitalization and balanced allocation. *Clim. Dev.*, **8**, 105–109,
36 doi:10.1080/17565529.2015.1040368.
- 37 Friis, C., and J. Ø. Nielsen, 2016: Small-scale land acquisitions, large-scale implications: Exploring
38 the case of Chinese banana investments in Northern Laos. *Land use policy*, **57**, 117–129,
39 doi:10.1016/j.landusepol.2016.05.028.
- 40 Fritsche, U., and Coauthors, 2017a: *GLOBAL LAND OUTLOOK WORKING PAPER: ENERGY AND*
41 *LAND USE*. Bonn, 60 pp. [https://knowledge.unccd.int/sites/default/files/2018-06/2-](https://knowledge.unccd.int/sites/default/files/2018-06/2-Fritsche%2Bet%2Bal%2B%282017%29%2BEnergy%2Band%2BLand%2BUse%2B-%2BGLO%2Bpaper-corr.pdf)
42 [Fritsche%2Bet%2Bal%2B%282017%29%2BEnergy%2Band%2BLand%2BUse%2B-](https://knowledge.unccd.int/sites/default/files/2018-06/2-Fritsche%2Bet%2Bal%2B%282017%29%2BEnergy%2Band%2BLand%2BUse%2B-%2BGLO%2Bpaper-corr.pdf)
43 [%2BGLO%2Bpaper-corr.pdf](https://knowledge.unccd.int/sites/default/files/2018-06/2-Fritsche%2Bet%2Bal%2B%282017%29%2BEnergy%2Band%2BLand%2BUse%2B-%2BGLO%2Bpaper-corr.pdf) (Accessed April 3, 2019).
- 44 ———, and Coauthors, 2017b: *GLOBAL LAND OUTLOOK WORKING PAPER: ENERGY AND LAND*
45 *USE*. Bonn, 60 pp.
- 46 Fuerth, L. S., 2009: Operationalizing Anticipatory Governance. *Prism*.
- 47 ———, and E. M. H. Faber, 2013: Anticipatory governance: Winning the future. *Futurist*.

- 1 Fujimori, S., T. Hasegawa, J. Rogelj, X. Su, P. Havlik, V. Krey, K. Takahashi, and K. Riahi, 2018a:
2 Inclusive climate change mitigation and food security policy under 1.5 °C climate goal. *Environ.*
3 *Res. Lett.*, **13**, 074033, doi:10.1088/1748-9326/aad0f7. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997)
4 [9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997) (Accessed April 10,
5 2019).
- 6 ———, ———, ———, ———, ———, ———, ———, and ———, 2018b: Inclusive climate change mitigation and
7 food security policy under 1.5 °C climate goal. *Environ. Res. Lett.*, **13**, 074033,
8 doi:10.1088/1748-9326/aad0f7. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997)
9 [9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997) (Accessed April 23,
10 2019).
- 11 Fuller, T., and M. Qingwen, 2013: Understanding Agricultural Heritage Sites as Complex Adaptive
12 Systems: The Challenge of Complexity. <https://doi.org/10.5814/j.issn.1674-764x.2013.03.002>,
13 **4**, 195–201, doi:10.5814/J.ISSN.1674-764X.2013.03.002. [https://bioone.org/journals/Journal-of-](https://bioone.org/journals/Journal-of-Resources-and-Ecology/volume-4/issue-3/j.issn.1674-764x.2013.03.002/Understanding-Agricultural-Heritage-Sites-as-Complex-Adaptive-Systems--The/10.5814/j.issn.1674-764x.2013.03.002.short)
14 [Resources-and-Ecology/volume-4/issue-3/j.issn.1674-764x.2013.03.002/Understanding-](https://bioone.org/journals/Journal-of-Resources-and-Ecology/volume-4/issue-3/j.issn.1674-764x.2013.03.002/Understanding-Agricultural-Heritage-Sites-as-Complex-Adaptive-Systems--The/10.5814/j.issn.1674-764x.2013.03.002.short)
15 [Agricultural-Heritage-Sites-as-Complex-Adaptive-Systems--The/10.5814/j.issn.1674-](https://bioone.org/journals/Journal-of-Resources-and-Ecology/volume-4/issue-3/j.issn.1674-764x.2013.03.002/Understanding-Agricultural-Heritage-Sites-as-Complex-Adaptive-Systems--The/10.5814/j.issn.1674-764x.2013.03.002.short)
16 [764x.2013.03.002.short](https://bioone.org/journals/Journal-of-Resources-and-Ecology/volume-4/issue-3/j.issn.1674-764x.2013.03.002/Understanding-Agricultural-Heritage-Sites-as-Complex-Adaptive-Systems--The/10.5814/j.issn.1674-764x.2013.03.002.short) (Accessed April 8, 2019).
- 17 Fung, A., 2015: Putting the Public Back into Governance: The Challenges of Citizen Participation and
18 Its Future. *Public Adm. Rev.*, **75**, 513–522, doi:10.1111/puar.12361.
19 <http://doi.wiley.com/10.1111/puar.12361> (Accessed October 2, 2018).
- 20 Furtado, F., 2018: A construção da natureza e a natureza da construção: políticas de incentivo aos
21 serviços ambientais no Acre e no Mato Grosso. *Estud. Soc. e Agric.*, 123–147.
22 <https://revistaesa.com/ojs/index.php/esa/article/view/1152> (Accessed March 9, 2019).
- 23 Fuso Nerini, F., C. Ray, and Y. Boulkaid, 2017: The cost of cooking a meal. The case of Nyeri
24 County, Kenya. *Environ. Res. Lett.*, **12**, 065007, doi:10.1088/1748-9326/aa6fd0.
25 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=6/a=065007?key=crossref.1232d9fca5f5811880c57142e5e84687)
26 [9326/12/i=6/a=065007?key=crossref.1232d9fca5f5811880c57142e5e84687](http://stacks.iop.org/1748-9326/12/i=6/a=065007?key=crossref.1232d9fca5f5811880c57142e5e84687) (Accessed May 23,
27 2018).
- 28 ———, and Coauthors, 2018: Mapping synergies and trade-offs between energy and the Sustainable
29 Development Goals. *Nat. Energy*, **3**, 10–15, doi:10.1038/s41560-017-0036-5.
30 <http://www.nature.com/articles/s41560-017-0036-5> (Accessed May 23, 2018).
- 31 Fuss, S., and Coauthors, 2014: Betting on negative emissions. *Nat. Clim. Chang.*, **4**, 850–853,
32 doi:10.1038/nclimate2392. <http://www.nature.com/doi/10.1038/nclimate2392> (Accessed
33 December 13, 2017).
- 34 ———, and Coauthors, 2018: Negative emissions—Part 2: Costs, potentials and side effects. *Environ.*
35 *Res. Lett.*, **13**, 063002, doi:10.1088/1748-9326/aabf9f.
- 36 Fussell, E., L. M. Hunter, and C. L. Gray, 2014: Measuring the environmental dimensions of human
37 migration: The demographer’s toolkit. *Glob. Environ. Chang.*,
38 doi:10.1016/j.gloenvcha.2014.07.001.
- 39 Fyson, C., and L. Jeffery, 2018: Examining treatment of the LULUCF sector in the NDCs. *20th EGU*
40 *Gen. Assem. EGU2018, Proc. from Conf. held 4-13 April. 2018 Vienna, Austria, p.16542, 20*,
41 16542. <http://adsabs.harvard.edu.ezp.sub.su.se/abs/2018EGUGA..2016542F> (Accessed October
42 15, 2018).
- 43 Gabay, M., and M. Alam, 2017: Community forestry and its mitigation potential in the Anthropocene:
44 The importance of land tenure governance and the threat of privatization. *For. Policy Econ.*, **79**,
45 26–35, doi:10.1016/j.forpol.2017.01.011.
- 46 Gallagher, J., 2014: Learning about an Infrequent Event: Evidence from Flood Insurance Take-Up in
47 the United States. *Am. Econ. J. Appl. Econ.*, **6**, 206–233, doi:10.1257/app.6.3.206.
48 <http://pubs.aeaweb.org/doi/10.1257/app.6.3.206> (Accessed April 3, 2019).
- 49 Gallina, V., S. Torresan, A. Critto, A. Sperotto, T. Glade, and A. Marcomini, 2016: A review of

- 1 multi-risk methodologies for natural hazards: Consequences and challenges for a climate change
2 impact assessment. *J. Environ. Manage.*, **168**, 123–132, doi:10.1016/j.jenvman.2015.11.011.
- 3 Gan, J., A. Jarrett, and C. J. Gaither, 2014: Wildfire risk adaptation: propensity of forestland owners
4 to purchase wildfire insurance in the southern United States. *Can. J. For. Res.*, **44**, 1376–1382,
5 doi:10.1139/cjfr-2014-0301. <http://www.nrcresearchpress.com/doi/10.1139/cjfr-2014-0301>
6 (Accessed May 16, 2018).
- 7 Gandenberger, C., M. Bodenheimer, J. Schleich, R. Orzanna, and L. Macht, 2016: Factors driving
8 international technology transfer: empirical insights from a CDM project survey. *Clim. Policy*,
9 **16**, 1065–1084, doi:10.1080/14693062.2015.1069176.
10 <https://www.tandfonline.com/doi/full/10.1080/14693062.2015.1069176> (Accessed March 17,
11 2019).
- 12 Garcia, C., and C. J. Fearnley, 2012: Evaluating critical links in early warning systems for natural
13 hazards. *Environmental Hazards*.
- 14 Gardner, T. A., and Coauthors, 2018a: Transparency and sustainability in global commodity supply
15 chains. doi:10.1016/j.worlddev.2018.05.025. <https://doi.org/10.1016/j.worlddev.2018.05.025>
16 (Accessed April 7, 2019).
- 17 —, and Coauthors, 2018b: Transparency and sustainability in global commodity supply chains.
18 doi:10.1016/j.worlddev.2018.05.025.
- 19 Garnett, S. T., and Coauthors, 2018: A spatial overview of the global importance of Indigenous lands
20 for conservation. *Nat. Sustain.*, **1**, 369–374, doi:10.1038/s41893-018-0100-6.
21 <http://www.nature.com/articles/s41893-018-0100-6> (Accessed April 11, 2019).
- 22 Garnett, T., and Coauthors, 2013: Sustainable Intensification in Agriculture: Premises and Policies.
23 *Science (80-.)*, **341**.
- 24 Garrett, R. D., and Coauthors, 2019: Criteria for effective zero-deforestation commitments.
25 doi:10.1016/j.gloenvcha.2018.11.003. <http://creativecommons.org/licenses/by-nc-nd/4.0/>
26 (Accessed April 7, 2019).
- 27 Gasparatos, A., and Coauthors, 2018a: Mechanisms and indicators for assessing the impact of biofuel
28 feedstock production on ecosystem services. *Biomass and Bioenergy*, **114**, 157–173,
29 doi:10.1016/j.biombioe.2018.01.024.
30 <https://linkinghub.elsevier.com/retrieve/pii/S0961953418300308> (Accessed April 13, 2019).
- 31 —, and Coauthors, 2018b: Survey of local impacts of biofuel crop production and adoption of
32 ethanol stoves in southern Africa. *Sci. Data*, **5**, 180186, doi:10.1038/sdata.2018.186.
33 <http://www.nature.com/articles/sdata2018186> (Accessed April 13, 2019).
- 34 Gasparatos, A., C. Romeu-Dalmau, G. von Maltitz, F. X. Johnson, C. B. Jumbe, P. Stromberg, and K.
35 Willis, 2018c: Using an ecosystem services perspective to assess biofuel sustainability. *Biomass
36 and Bioenergy*, **114**, 1–7, doi:10.1016/j.biombioe.2018.01.025.
37 <https://linkinghub.elsevier.com/retrieve/pii/S096195341830031X> (Accessed April 13, 2019).
- 38 Gazol, A., J. J. Camarero, J. J. Jiménez, D. Moret-Fernández, M. V. López, G. Sangüesa-Barreda, and
39 J. M. Igual, 2018: Beneath the canopy: Linking drought-induced forest die off and changes in
40 soil properties. *For. Ecol. Manage.*, doi:10.1016/j.foreco.2018.04.028.
- 41 GBEP, 2017: *The Global Bioenergy Partnership A Global Commitment to Bioenergy*. 4 pp.
42 [http://www.globalbioenergy.org/fileadmin/user_upload/gbep/docs/ENGLISH_Background_note
43 _GBEP_Setember_2016_FINAL.pdf](http://www.globalbioenergy.org/fileadmin/user_upload/gbep/docs/ENGLISH_Background_note_GBEP_Setember_2016_FINAL.pdf) (Accessed April 8, 2019).
- 44 Geddes, A., T. S. Schmidt, and B. Steffen, 2018: The multiple roles of state investment banks in low-
45 carbon energy finance: An analysis of Australia, the UK and Germany. *Energy Policy*, **115**,
46 158–170, doi:10.1016/j.enpol.2018.01.009.
- 47 Geden, O., G. P. Peters, and V. Scott, 2019: Targeting carbon dioxide removal in the European Union.
48 *Clim. Policy*, **19**, 487–494, doi:10.1080/14693062.2018.1536600.

- 1 <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1536600> (Accessed March 30,
2 2019).
- 3 van der Geest, K., and K. Warner, 2014: Vulnerability, Coping and Loss and Damage from Climate
4 Events. *Hazards, Risks and, Disasters in Society*.
- 5 Geisler, C., and B. Currens, 2017: Impediments to inland resettlement under conditions of accelerated
6 sea level rise. *Land use policy*, doi:10.1016/j.landusepol.2017.03.029.
- 7 German, L., and G. Schoneveld, 2012: A review of social sustainability considerations among EU-
8 approved voluntary schemes for biofuels, with implications for rural livelihoods. *Energy Policy*,
9 **51**, 765–778, doi:10.1016/J.ENPOL.2012.09.022. [https://www-sciencedirect-](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S0301421512007975)
10 [com.ezp.sub.su.se/science/article/pii/S0301421512007975](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S0301421512007975) (Accessed May 23, 2018).
- 11 Gersonius, B., R. Ashley, A. Pathirana, and C. Zevenbergen, 2013: Climate change uncertainty:
12 building flexibility into water and flood risk infrastructure. *Clim. Change*, **116**, 411–423.
- 13 Gheewala, S. H., G. Berndes, and G. Jewitt, 2011: The bioenergy and water nexus. *Biofuels, Bioprod.*
14 *Biorefining*, **5**, 353–360.
- 15 Ghilardi, A., and Coauthors, 2016a: Spatiotemporal modeling of fuelwood environmental impacts:
16 Towards improved accounting for non-renewable biomass. *Environ. Model. Softw.*, **82**, 241–
17 254, doi:10.1016/j.envsoft.2016.04.023.
18 <https://linkinghub.elsevier.com/retrieve/pii/S1364815216301189> (Accessed April 2, 2019).
- 19 —, and Coauthors, 2016b: Spatiotemporal modeling of fuelwood environmental impacts: Towards
20 improved accounting for non-renewable biomass. *Environ. Model. Softw.*, **82**, 241–254,
21 doi:10.1016/j.envsoft.2016.04.023.
- 22 Ghilardi, A., A. Tarter, and R. Bailis, 2018: Potential environmental benefits from woodfuel
23 transitions in Haiti: Geospatial scenarios to 2027. *Environ. Res. Lett.*, **13**, 035007,
24 doi:10.1088/1748-9326/aaa846. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=3/a=035007?key=crossref.a43a541bcd114bd335c7546a250389dc)
25 [9326/13/i=3/a=035007?key=crossref.a43a541bcd114bd335c7546a250389dc](http://stacks.iop.org/1748-9326/13/i=3/a=035007?key=crossref.a43a541bcd114bd335c7546a250389dc) (Accessed April 2,
26 2019).
- 27 Ghosh, A., S. Schmidt, T. Fickert, and M. Nüsser, 2015: The Indian Sundarban mangrove forests:
28 history, utilization, conservation strategies and local perception. *Diversity*, **7**, 149–169.
- 29 Ghosh, S., H. Vittal, T. Sharma, S. Karmakar, K. S. Kasiviswanathan, Y. Dhanesh, K. P. Sudheer, and
30 S. S. Gunthe, 2016: Indian Summer Monsoon Rainfall: Implications of Contrasting Trends in the
31 Spatial Variability of Means and Extremes. *PLoS One*, **11**, e0158670,
32 doi:10.1371/journal.pone.0158670. <http://dx.plos.org/10.1371/journal.pone.0158670> (Accessed
33 May 18, 2018).
- 34 Gibson, L., E. N. Wilman, and W. F. Laurance, 2017: How Green is ‘Green’ Energy? *Trends Ecol.*
35 *Evol.*, **32**, 922–935.
- 36 Giessen, L., S. Burns, M. A. K. Sahide, and A. Wibowo, 2016a: From governance to government: The
37 strengthened role of state bureaucracies in forest and agricultural certification. *Policy Soc.*, **35**,
38 71–89, doi:10.1016/j.polsoc.2016.02.001.
39 <https://www.tandfonline.com/doi/full/10.1016/j.polsoc.2016.02.001> (Accessed April 3, 2019).
- 40 —, —, —, and —, 2016b: From governance to government: The strengthened role of state
41 bureaucracies in forest and agricultural certification. *Policy Soc.*, **35**, 71–89,
42 doi:10.1016/j.polsoc.2016.02.001.
- 43 Gilbert, C. L., 2010: How to understand high food prices. *J. Agric. Econ.*, doi:10.1111/j.1477-
44 9552.2010.00248.x.
- 45 Gill, A. B., 2005: Offshore renewable energy: ecological implications of generating electricity in the
46 coastal zone. *J. Appl. Ecol.*, **42**, 605–615.
- 47 Girard, C., M. Pulido-Velazquez, J.-D. Rinaudo, C. Pagé, and Y. Caballero, 2015: Integrating top–

- 1 down and bottom-up approaches to design global change adaptation at the river basin scale.
2 *Glob. Environ. Chang.*, **34**, 132–146, doi:<https://doi.org/10.1016/j.gloenvcha.2015.07.002>.
- 3 Girma, H. M., R. M. Hassan, and G. Hertzler, 2012: Forest conservation versus conversion under
4 uncertain market and environmental forest benefits in Ethiopia: The case of Sheka forest. *For.*
5 *Policy Econ.*, doi:10.1016/j.forpol.2012.01.001.
- 6 Gitau, J. K., and Coauthors, 2019: Implications on Livelihoods and the Environment of Uptake of
7 Gasifier Cook Stoves among Kenya's Rural Households. *Appl. Sci.*, **9**, 1205,
8 doi:10.3390/app9061205. <https://www.mdpi.com/2076-3417/9/6/1205> (Accessed April 14,
9 2019).
- 10 Glachant, M., and A. Dechezleprêtre, 2017: What role for climate negotiations on technology
11 transfer? *Clim. Policy*, **17**, 962–981, doi:10.1080/14693062.2016.1222257.
12 <https://www.tandfonline.com/doi/full/10.1080/14693062.2016.1222257> (Accessed May 22,
13 2018).
- 14 Glauben, T., T. Herzfeld, S. Rozelle, and X. Wang, 2012: Persistent Poverty in Rural China: Where,
15 Why, and How to Escape? *World Dev.*, **40**, 784–795,
16 doi:<https://doi.org/10.1016/j.worlddev.2011.09.023>.
17 <http://www.sciencedirect.com/science/article/pii/S0305750X11002452>.
- 18 Gleick, P. H., 2014: Water, drought, climate change, and conflict in Syria. *Weather. Clim. Soc.*, **6**,
19 331–340.
- 20 Glemarec, Y., 2017: Addressing the gender differentiated investment risks to climate-smart
21 agriculture. *AIMS Agric. Food*, **2**, 56–74, doi:10.3934/agrfood.2017.1.56.
22 <http://www.aimspress.com/article/10.3934/agrfood.2017.1.56> (Accessed April 2, 2019).
- 23 Go, A. W., A. T. Conag, R. M. B. Igdon, A. S. Toledo, and J. S. Malila, 2019a: Potentials of
24 agricultural and agro-industrial crop residues for the displacement of fossil fuels: A Philippine
25 context. *Energy Strateg. Rev.*, **23**, 100–113, doi:10.1016/j.esr.2018.12.010. [https://www-](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S2211467X18301226)
26 [sciencedirect-com.ezp.sub.su.se/science/article/pii/S2211467X18301226](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S2211467X18301226) (Accessed April 14,
27 2019).
- 28 —, —, —, —, and —, 2019b: Potentials of agricultural and agro-industrial crop residues
29 for the displacement of fossil fuels: A Philippine context. *Energy Strateg. Rev.*, **23**, 100–113,
30 doi:10.1016/j.esr.2018.12.010.
- 31 Godar, J., and T. Gardner, 2019: Trade and Land-Use Telecouplings. *Telecoupling*, Springer
32 International Publishing, Cham, 149–175 [http://link.springer.com/10.1007/978-3-030-11105-](http://link.springer.com/10.1007/978-3-030-11105-2_8)
33 [2_8](http://link.springer.com/10.1007/978-3-030-11105-2_8) (Accessed April 8, 2019).
- 34 —, U. M. Persson, E. J. Tizado, and P. Meyfroidt, 2015: Methodological and Ideological Options
35 Towards more accurate and policy relevant footprint analyses: Tracing fine-scale socio-
36 environmental impacts of production to consumption. *Ecol. Econ.*, **112**, 25–35,
37 doi:10.1016/j.ecolecon.2015.02.003. <http://dx.doi.org/10.1016/j.ecolecon.2015.02.003>
38 (Accessed April 8, 2019).
- 39 —, C. Suavet, T. A. Gardner, E. Dawkins, and P. Meyfroidt, 2016: Balancing detail and scale in
40 assessing transparency to improve the governance of agricultural commodity supply chains.
41 *Environ. Res. Lett.*, **11**, doi:10.1088/1748-9326/11/3/035015. [https://iopscience-iop-](https://iopscience-iop-org.ezp.sub.su.se/article/10.1088/1748-9326/11/3/035015/pdf)
42 [org.ezp.sub.su.se/article/10.1088/1748-9326/11/3/035015/pdf](https://iopscience-iop-org.ezp.sub.su.se/article/10.1088/1748-9326/11/3/035015/pdf) (Accessed April 8, 2019).
- 43 Goetz, A., 2013: Private Governance and Land Grabbing: The Equator Principles and the Roundtable
44 on Sustainable Biofuels. *GLOBALIZATIONS*, **10**, 199–204,
45 doi:10.1080/14747731.2013.760949.
- 46 Goetz, S. J., M. Hansen, R. A. Houghton, W. Walker, N. Laporte, and J. Busch, 2015: Measurement
47 and monitoring needs, capabilities and potential for addressing reduced emissions from
48 deforestation and forest degradation under REDD+. *Environ. Res. Lett.*, **10**, 123001,
49 doi:10.1088/1748-9326/10/12/123001. [Subject to Copy-editing](http://stacks.iop.org/1748-</p></div><div data-bbox=)

- 1 9326/10/i=12/a=123001?key=crossref.0c01a5150ad68eef41d5cef761c1fe70 (Accessed October
2 15, 2018).
- 3 Goh, A. H. X., 2012: A Literature Review of the Gender-differentiated Impacts of Climate Change
4 On Women's and Men's Assets and Well-being in Developing Countries. *Int. Food Policy Res.*
5 *Inst.*, **CAPRI Work**, doi:10.2499/CAPRIWP106.
- 6 Golay, C., and I. Biglino, 2013: Human Rights Responses to Land Grabbing: a right to food
7 perspective. *Third World Q.*, **34**, 1630–1650, doi:10.1080/01436597.2013.843853.
- 8 Gold Standard, Gold Standard for the Global Goals. 2018,. [https://www.goldstandard.org/our-](https://www.goldstandard.org/our-work/what-we-do)
9 [work/what-we-do](https://www.goldstandard.org/our-work/what-we-do) (Accessed April 8, 2019).
- 10 Goldemberg, J., and S. Teixeira Coelho, 2004: Renewable energy-traditional biomass vs. modern
11 biomass. *Energy Policy*, **32**, 711–714, doi:10.1016/S0301-4215(02)00340-3.
12 [https://pdf.sciencedirectassets.com/271097/1-s2.0-S0301421500X02716/1-s2.0-](https://pdf.sciencedirectassets.com/271097/1-s2.0-S0301421500X02716/1-s2.0-S0301421502003403/main.pdf?x-amz-security-token=AgoJb3JpZ2luX2VjEBAaCXVzLWVhc3QtMSJHMEUCIQD9GM6XbFe5Y1cQOB38lu)
13 [S0301421502003403/main.pdf?x-amz-security-](https://pdf.sciencedirectassets.com/271097/1-s2.0-S0301421502003403/main.pdf?x-amz-security-token=AgoJb3JpZ2luX2VjEBAaCXVzLWVhc3QtMSJHMEUCIQD9GM6XbFe5Y1cQOB38lu)
14 [token=AgoJb3JpZ2luX2VjEBAaCXVzLWVhc3QtMSJHMEUCIQD9GM6XbFe5Y1cQOB38lu](https://pdf.sciencedirectassets.com/271097/1-s2.0-S0301421502003403/main.pdf?x-amz-security-token=AgoJb3JpZ2luX2VjEBAaCXVzLWVhc3QtMSJHMEUCIQD9GM6XbFe5Y1cQOB38lu)
15 [x7OzaaTw6pMtKmVfeW9bqJtAlgYV26FhAoJEoe1kwYnmybc66qvs%2FKx1NWxYkVOBi9](https://pdf.sciencedirectassets.com/271097/1-s2.0-S0301421502003403/main.pdf?x-amz-security-token=AgoJb3JpZ2luX2VjEBAaCXVzLWVhc3QtMSJHMEUCIQD9GM6XbFe5Y1cQOB38lu)
16 [ZU](https://pdf.sciencedirectassets.com/271097/1-s2.0-S0301421502003403/main.pdf?x-amz-security-token=AgoJb3JpZ2luX2VjEBAaCXVzLWVhc3QtMSJHMEUCIQD9GM6XbFe5Y1cQOB38lu) (Accessed April 14, 2019).
- 17 —, J. Martinez-Gomez, A. Sagar, and K. R. Smith, 2018a: Household air pollution, health, and
18 climate change: cleaning the air. *Environ. Res. Lett.*, **13**, 030201, doi:10.1088/1748-
19 9326/aaa49d. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=3/a=030201?key=crossref.1c1d7c558768fe35c23196a5c80296b1)
20 [9326/13/i=3/a=030201?key=crossref.1c1d7c558768fe35c23196a5c80296b1](http://stacks.iop.org/1748-9326/13/i=3/a=030201?key=crossref.1c1d7c558768fe35c23196a5c80296b1) (Accessed April 14,
21 2019).
- 22 —, —, —, and —, 2018b: Household air pollution, health, and climate change: cleaning the
23 air. *Environ. Res. Lett.*, **13**, 030201, doi:10.1088/1748-9326/aaa49d.
- 24 Gómez-Baggethun, E., S. Míngorría, V. Reyes-García, L. Calvet, and C. Montes, 2010: Traditional
25 Ecological Knowledge Trends in the Transition to a Market Economy: Empirical Study in the
26 Doñana Natural Areas. *Conserv. Biol.*, **24**, 721–729, doi:10.1111/j.1523-1739.2009.01401.x.
27 <http://doi.wiley.com/10.1111/j.1523-1739.2009.01401.x> (Accessed April 6, 2019).
- 28 Gopakumar, G., 2014: *Transforming Urban Water Supplies in India*. Routledge,.
- 29 Gordon, S. M., 2016: The Foreign Corrupt Practices Act: Prosecute Corruption and End Transnational
30 Illegal Logging. *Bost. Coll. Environ. Aff. Law Rev.*, **43**.
31 <https://heinonline.org/HOL/Page?handle=hein.journals/bcenv43&id=122&div=8&collection=journals>
32 [journals](https://heinonline.org/HOL/Page?handle=hein.journals/bcenv43&id=122&div=8&collection=journals) (Accessed November 2, 2018).
- 33 Götz, L., T. Glauben, and B. Brümmer, 2013: Wheat export restrictions and domestic market effects
34 in Russia and Ukraine during the food crisis. *Food Policy*, doi:10.1016/j.foodpol.2012.12.001.
- 35 Goudie, A. S., 2014: Desert dust and human health disorders. *Environ. Int.*, **63**, 101–113,
36 doi:10.1016/J.ENVINT.2013.10.011.
37 <https://www.sciencedirect.com/science/article/pii/S0160412013002262?via%3Dihub> (Accessed
38 April 14, 2019).
- 39 Goulder, L. H., and R. C. Williams, 2012: *The Choice of Discount Rate for Climate Change Policy*
40 *Evaluation*.
- 41 Government of France, 2019: Ending deforestation caused by importing unsustainable products |
42 Gouvernement.fr. *Website*,. [https://www.gouvernement.fr/en/ending-deforestation-caused-by-](https://www.gouvernement.fr/en/ending-deforestation-caused-by-importing-unsustainable-products)
43 [importing-unsustainable-products](https://www.gouvernement.fr/en/ending-deforestation-caused-by-importing-unsustainable-products) (Accessed April 8, 2019).
- 44 Government of India, *National Mission for a Green India. Under The National Action Plan on*
45 *Climate Change*. New Delhi, http://www.moef.gov.in/sites/default/files/GIM_Mission
46 [Document-1.pdf](http://www.moef.gov.in/sites/default/files/GIM_Mission) (Accessed April 14, 2019).
- 47 —, 2012: *Guidelines for Preparation of State Action Plan for Bustards' Recovery Programme*.
48 <http://envfor.nic.in/assets/wl-2072012.pdf> (Accessed October 30, 2018).

- 1 Government Office for Science, 2011: Migration and Global Environmental Change: Future
2 Challenges and Opportunities. *Gov. Off. Sci. - Foresight*, doi:10.1021/jp801592w.
- 3 Van de Graaf, T., 2017: Is OPEC dead? Oil exporters, the Paris agreement and the transition to a post-
4 carbon world. *Energy Res. Soc. Sci.*, **23**, 182–188.
- 5 Graham, L. J., R. H. Haines-Young, and R. Field, 2015: Using citizen science data for conservation
6 planning: Methods for quality control and downscaling for use in stochastic patch occupancy
7 modelling. *Biol. Conserv.*, **192**, 65–73, doi:10.1016/j.biocon.2015.09.002.
- 8 Grainger, A., 2015: Is Land Degradation Neutrality feasible in dry areas? *J. Arid Environ.*, **112**, 14–
9 24, doi:10.1016/j.jaridenv.2014.05.014.
- 10 Gray, S., and Coauthors, 2017: Combining participatory modelling and citizen science to support
11 volunteer conservation action. *Biol. Conserv.*, **208**, 76–86, doi:10.1016/J.BIOCON.2016.07.037.
12 <http://www.sciencedirect.com/science/article/pii/S0006320716303020> (Accessed December 11,
13 2017).
- 14 Greatrex, H., and Coauthors, 2015: Scaling up index insurance for smallholder farmers: Recent
15 evidence and insights. *CCAFS Rep.*, **14**, 1–32, doi:1904-9005.
- 16 Grecchi, R. C., Q. H. J. Gwyn, G. B. Bénié, A. R. Formaggio, and F. C. Fahl, 2014: Land use and
17 land cover changes in the Brazilian Cerrado: A multidisciplinary approach to assess the impacts
18 of agricultural expansion. *Appl. Geogr.*, **55**, 300–312, doi:10.1016/J.APGEOG.2014.09.014.
19 <https://www.sciencedirect.com/science/article/abs/pii/S0143622814002240> (Accessed April 23,
20 2019).
- 21 Green, D., and G. Raygorodetsky, 2010: Indigenous knowledge of a changing climate. *Clim. Change*,
22 **100**, 239–242.
- 23 Di Gregorio, M., and Coauthors, 2017: Climate policy integration in the land use sector: Mitigation,
24 adaptation and sustainable development linkages. *Environ. Sci. Policy*, **67**, 35–43,
25 doi:10.1016/j.envsci.2016.11.004.
- 26 Greiner, R., and D. Gregg, 2011: Farmers' intrinsic motivations, barriers to the adoption of
27 conservation practices and effectiveness of policy instruments: Empirical evidence from northern
28 Australia. *Land use policy*, **28**, 257–265.
- 29 Griggs, D., and Coauthors, 2013: Sustainable development goals for people and planet. *Nature*, **495**,
30 305. <http://dx.doi.org/10.1038/495305a>.
- 31 ———, and Coauthors, 2014: An integrated framework for sustainable development goals. *Ecol. Soc.*,
32 **19**, art49-art49, doi:10.5751/ES-07082-190449.
33 [http://www.scopus.com/inward/record.url?eid=2-s2.0-
34 84924356905&partnerID=tZOtx3y1%5Cnhttp://www.ecologyandsociety.org/vol19/iss4/art49/](http://www.scopus.com/inward/record.url?eid=2-s2.0-84924356905&partnerID=tZOtx3y1%5Cnhttp://www.ecologyandsociety.org/vol19/iss4/art49/).
- 35 De Groot, R. S., R. Alkemade, L. Braat, L. Hein, and L. Willemen, 2010: Challenges in integrating
36 the concept of ecosystem services and values in landscape planning, management and decision
37 making. *Ecol. Complex.*, **7**, 260–272.
- 38 Grosjean, G., and Coauthors, 2018: Options to overcome the barriers to pricing European agricultural
39 emissions. *Clim. Policy*, **18**, 151–169, doi:10.1080/14693062.2016.1258630.
40 <https://doi.org/10.1080/14693062.2016.1258630>.
- 41 Del Grosso, S., P. Smith, M. Galdos, A. Hastings, and W. Parton, 2014: Sustainable energy crop
42 production. *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2014.07.007.
- 43 Grumbine, R. E., and M. K. Pandit, 2013: Threats from India's Himalaya dams. *Science (80-.)*, **339**,
44 36–37.
- 45 Grzymala-Busse, A., 2010: The best laid plans: The impact of informal rules on formal institutions in
46 transitional regimes. *Stud. Comp. Int. Dev.*, **45**, 311–333, doi:10.1007/s12116-010-9071-y.
- 47 Gu, Y., and B. K. Wylie, 2017: Mapping marginal croplands suitable for cellulosic feedstock crops in

- 1 the Great Plains, United States. *GCB Bioenergy*, doi:10.1111/gcbb.12388.
- 2 Guerry, A. D., and Coauthors, 2015: Natural capital and ecosystem services informing decisions:
3 From promise to practice. *Proc. Natl. Acad. Sci.*, **112**, 7348–7355.
- 4 Gunderson, L. H., and C. . Holling, 2001: Panarchy: Understanding transformations. *Hum. Nat. Syst.*,
5 507.
- 6 Gupta, H., and L. C. Dube, 2018: Addressing biodiversity in climate change discourse: Paris
7 mechanisms hold more promise. *Int. For. Rev.*, **20**, 104–114,
8 doi:10.1505/146554818822824282.
9 <http://www.ingentaconnect.com/content/10.1505/146554818822824282> (Accessed March 17,
10 2019).
- 11 Gupta, J., 2014: *The History of Global Climate Governance*. 1-244 pp.
- 12 —, and C. Vegelin, 2016: Sustainable development goals and inclusive development. *Int. Environ.*
13 *Agreements Polit. Law Econ.*, doi:10.1007/s10784-016-9323-z.
- 14 —, and N. Pouw, 2017: Towards a trans-disciplinary conceptualization of inclusive development.
15 *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2017.03.004.
- 16 —, C. Termeer, J. Klostermann, S. Meijerink, M. van den Brink, P. Jong, S. Nooteboom, and E.
17 Bergsma, 2010: The Adaptive Capacity Wheel: A method to assess the inherent characteristics
18 of institutions to enable the adaptive capacity of society. *Environ. Sci. Policy*, **13**, 459–471,
19 doi:10.1016/j.envsci.2010.05.006.
- 20 Gupta, J., N. van der Grijp, and O. Kuik, 2013a: *Climate Change, Forests, and REDD Lessons for*
21 *Institutional Design*. J. Gupta, N. Van der Grijp, and O. Kuik, Eds. Routledge, London and New
22 York,.
- 23 Gupta, J., C. Pahl-Wostl, and R. Zondervan, 2013b: “Glocal” water governance: a multi-level
24 challenge in the anthropocene. *Curr. Opin. Environ. Sustain.*, **5**, 573–580,
25 doi:10.1016/j.cosust.2013.09.003.
- 26 —, N. R. M. Pouw, and M. A. F. Ros-Tonen, 2015: Towards an Elaborated Theory of Inclusive
27 Development. *Eur. J. Dev. Res.*, doi:10.1057/ejdr.2015.30.
- 28 Gurung, A., and S. E. Oh, 2013a: Conversion of traditional biomass into modern bioenergy systems:
29 A review in context to improve the energy situation in Nepal. *Renew. Energy*,
30 doi:10.1016/j.renene.2012.06.021. <http://dx.doi.org/10.1016/j.renene.2012.06.021> (Accessed
31 April 14, 2019).
- 32 —, and —, 2013b: Conversion of traditional biomass into modern bioenergy systems: A review
33 in context to improve the energy situation in Nepal. *Renew. Energy*,
34 doi:10.1016/j.renene.2012.06.021.
- 35 Haasnoot, M., J. H. Kwakkel, W. E. Walker, and J. ter Maat, 2013: Dynamic adaptive policy
36 pathways: A method for crafting robust decisions for a deeply uncertain world. *Glob. Environ.*
37 *Chang.*, **23**, 485–498, doi:10.1016/j.gloenvcha.2012.12.006.
- 38 —, S. van 't Klooster, and J. van Alphen, 2018: Designing a monitoring system to detect signals to
39 adapt to uncertain climate change. *Glob. Environ. Chang.*, **52**, 273–285,
40 doi:https://doi.org/10.1016/j.gloenvcha.2018.08.003.
- 41 Hadarits, M., J. Pittman, D. Corkal, H. Hill, K. Bruce, and A. Howard, 2017: The interplay between
42 incremental, transitional, and transformational adaptation: a case study of Canadian agriculture.
43 *Reg. Environ. Chang.*, **17**, 1515–1525.
- 44 Haddeland, I., and Coauthors, 2014: Global water resources affected by human interventions and
45 climate change. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 3251–3256, doi:10.1073/pnas.1222475110.
46 <http://www.ncbi.nlm.nih.gov/pubmed/24344275> (Accessed April 11, 2019).
- 47 Haites, E., 2018a: Carbon taxes and greenhouse gas emissions trading systems: what have we

- 1 learned? *Clim. Policy*, **18**, 955–966, doi:10.1080/14693062.2018.1492897.
2 <https://doi.org/10.1080/14693062.2018.1492897>.
- 3 —, 2018b: Carbon taxes and greenhouse gas emissions trading systems: what have we learned?
4 *Clim. Policy*, **18**, 955–966, doi:10.1080/14693062.2018.1492897.
- 5 Hajjar, R., R. A. Kozak, H. El-Lakany, and J. L. Innes, 2013: Community forests for forest
6 communities: Integrating community-defined goals and practices in the design of forestry
7 initiatives. *Land use policy*, doi:10.1016/j.landusepol.2013.03.002.
- 8 Hall, R., M. Edelman, S. M. Borrás Jr., I. Scoones, B. White, and W. Wolford, 2015: Resistance,
9 acquiescence or incorporation? An introduction to landgrabbing and political reactions `from
10 below'. *J. Peasant Stud.*, **42**, 467–488, doi:10.1080/03066150.2015.1036746.
- 11 Hall, S. J., R. Hilborn, N. L. Andrew, and E. H. Allison, 2013: Innovations in capture fisheries are an
12 imperative for nutrition security in the developing world. *Proc. Natl. Acad. Sci.*, **110**, 8393–
13 8398.
- 14 Hallegatte, S., 2009: Strategies to adapt to an uncertain climate. *Glob. Environ. Chang.*, **19**, 240–247.
- 15 Hallegatte, S., A. Shah, R. J. Lempert, C. Brown, and S. Gill, 2012: *Investment decision making under*
16 *deep uncertainty--application to climate change*. W. Bank, Ed. Washington,.
- 17 Hallegatte, S., C. Green, R. J. Nicholls, and J. Corfee-Morlot, 2013: Future flood losses in major
18 coastal cities. *Nat. Clim. Chang.*, **3**, 802.
- 19 —, A. Vogt-Schilb, M. Bangalore, and J. Rozenberg, 2017: *Unbreakable: Building the Resilience*
20 *of the Poor in the Face of Natural Disasters*. Washington, DC,.
- 21 Haller, T., G. Acciaioli, and S. Rist, 2016: Constitutionality: Conditions for Crafting Local Ownership
22 of Institution-Building Processes. *Soc. Nat. Resour.*, doi:10.1080/08941920.2015.1041661.
- 23 Hanasaki, N., T. Inuzuka, S. Kanae, and T. Oki, 2010: An estimation of global virtual water flow and
24 sources of water withdrawal for major crops and livestock products using a global hydrological
25 model. *J. Hydrol.*, **384**, 232–244.
- 26 Hanasaki, N., and Coauthors, 2013a: A global water scarcity assessment under Shared Socio-
27 economic Pathways – Part 2: Water availability and scarcity. *Hydrol. Earth*
28 *Syst. Sci.*, **17**, 2393–2413, doi:10.5194/hess-17-2393-2013. [https://www.hydrol-earth-syst-](https://www.hydrol-earth-syst-sci.net/17/2393/2013/)
29 [sci.net/17/2393/2013/](https://www.hydrol-earth-syst-sci.net/17/2393/2013/) (Accessed April 11, 2019).
- 30 —, and Coauthors, 2013b: A global water scarcity assessment under Shared Socio-economic
31 Pathways – Part 2: Water availability and scarcity. *Hydrol. Earth Syst. Sci.*, **17**,
32 2393–2413, doi:10.5194/hess-17-2393-2013.
- 33 Handmer, J., and Coauthors, 2012: Changes in impacts of climate extremes: Human systems and
34 ecosystems. *Managing the Risks of Extreme Events and Disasters to Advance Climate Change*
35 *Adaptation: Special Report of the Intergovernmental Panel on Climate Change*.
- 36 Hanger, S., C. Haug, T. Lung, and L. M. Bouwer, 2015: Mainstreaming climate change in regional
37 development policy in Europe: five insights from the 2007–2013 programming period. *Reg.*
38 *Environ. Chang.*, **15**, 973–985, doi:10.1007/s10113-013-0549-9.
39 <http://link.springer.com/10.1007/s10113-013-0549-9> (Accessed December 30, 2017).
- 40 Hanjra, M. A., and M. Ejaz Qureshi, 2010: Global water crisis and future food security in an era of
41 climate change. *Food Policy*, **35**, 365–377, doi:10.1016/j.foodpol.2010.05.006.
42 <https://pdfs.semanticscholar.org/ca39/7d84f756252ca7a2345cb9583d0729cd9c12.pdf> (Accessed
43 May 22, 2018).
- 44 Hanson, S., R. Nicholls, N. Ranger, S. Hallegatte, J. Corfee-Morlot, C. Herweijer, and J. Chateau,
45 2011: A global ranking of port cities with high exposure to climate extremes. *Clim. Change*,
46 doi:10.1007/s10584-010-9977-4.
- 47 Härdle, W. K., and B. L. Cabrera, 2010: Calibrating CAT bonds for Mexican earthquakes. *J. Risk*

- 1 *Insur.*, **77**, 625–650, doi:10.1111/j.1539-6975.2010.01355.x.
- 2 Harmsworth, G., and S. Awatere, 2013: 2013. Indigenous Māori knowledge and perspectives of
3 ecosystems. *Ecosystem services in New Zealand – conditions and trends*, J. Dymond, Ed.,
4 Manaaki Whenua Press, Lincoln, New Zealand., p. 274 – 286.
5 [https://www.researchgate.net/publication/258423762_Harmsworth_GR_Awatere_S_2013_Indig](https://www.researchgate.net/publication/258423762_Harmsworth_GR_Awatere_S_2013_Indigenous_Maori_knowledge_and_perspectives_of_ecosystems_In_Dymond_JR_ed_Ecosystem_services_in_New_Zealand_-_conditions_and_trends_Manaki_Whenua_Press_Lincoln_New_Zealand)
6 [enous_Maori_knowledge_and_perspectives_of_ecosystems_In_Dymond_JR_ed_Ecosystem_ser](https://www.researchgate.net/publication/258423762_Harmsworth_GR_Awatere_S_2013_Indigenous_Maori_knowledge_and_perspectives_of_ecosystems_In_Dymond_JR_ed_Ecosystem_services_in_New_Zealand_-_conditions_and_trends_Manaki_Whenua_Press_Lincoln_New_Zealand)
7 [vices_in_New_Zealand_-](https://www.researchgate.net/publication/258423762_Harmsworth_GR_Awatere_S_2013_Indigenous_Maori_knowledge_and_perspectives_of_ecosystems_In_Dymond_JR_ed_Ecosystem_services_in_New_Zealand_-_conditions_and_trends_Manaki_Whenua_Press_Lincoln_New_Zealand)
8 [_conditions_and_trends_Manaki_Whenua_Press_Lincoln_New_Zealan](https://www.researchgate.net/publication/258423762_Harmsworth_GR_Awatere_S_2013_Indigenous_Maori_knowledge_and_perspectives_of_ecosystems_In_Dymond_JR_ed_Ecosystem_services_in_New_Zealand_-_conditions_and_trends_Manaki_Whenua_Press_Lincoln_New_Zealand) (Accessed April 24,
9 2019).
- 10 Harootunian, G., 2018: California: It's Complicated: Drought, Drinking Water and Drylands.
11 *Resilience: The Science of Adaptation to Climate Change*, K. Alverson and Z. Zommers, Eds.,
12 Elsevier, London.
- 13 Harris, E., 2013: Financing social protection floors: Considerations of fiscal space. *Int. Soc. Secur.*
14 *Rev.*, **66**, 111–143, doi:10.1111/issr.12021.
- 15 Harris, Z. M., R. Spake, and G. Taylor, 2015: Land use change to bioenergy: A meta-analysis of soil
16 carbon and GHG emissions. *Biomass and Bioenergy*, doi:10.1016/j.biombioe.2015.05.008.
- 17 Harrod, K. S., 2015: Ebola: history, treatment, and lessons from a new emerging pathogen. *Am. J.*
18 *Physiol. - Lung Cell. Mol. Physiol.*, **308**, L307–L313, doi:10.1152/ajplung.00354.2014.
- 19 Harvey, B., J. Ensor, L. Carlile, B. Garside, and Z. Patterson, 2012: Climate change communication
20 and social learning-Review and strategy development for CCAFS.
- 21 Harvey, C. A., and Coauthors, 2014a: Climate-Smart Landscapes: Opportunities and Challenges for
22 Integrating Adaptation and Mitigation in Tropical Agriculture. *Conserv. Lett.*, **7**, 77–90,
23 doi:10.1111/conl.12066. <http://doi.wiley.com/10.1111/conl.12066> (Accessed December 21,
24 2017).
- 25 —, Z. L. Rakotobe, N. S. Rao, R. Dave, H. Razafimahatratra, R. H. Rabarijohn, H. Rajaofara, and
26 J. L. MacKinnon, 2014b: Extreme vulnerability of smallholder farmers to agricultural risks and
27 climate change in Madagascar. *Philos. Trans. R. Soc. B Biol. Sci.*, doi:10.1098/rstb.2013.0089.
- 28 Harvey, M., and S. Pilgrim, 2011: The new competition for land: Food, energy, and climate change.
29 *Food Policy*, doi:10.1016/j.foodpol.2010.11.009.
- 30 Hasegawa, T., S. Fujimori, K. Takahashi, T. Yokohata, and T. Masui, 2016: Economic implications of
31 climate change impacts on human health through undernourishment. *Clim. Change*, **136**, 189–
32 202, doi:10.1007/s10584-016-1606-4. <http://dx.doi.org/10.1007/s10584-016-1606-4>.
- 33 —, and Coauthors, 2018a: Risk of increased food insecurity under stringent global climate change
34 mitigation policy. *Nat. Clim. Chang.*, **8**, 699–703, doi:10.1038/s41558-018-0230-x.
35 <http://www.nature.com/articles/s41558-018-0230-x> (Accessed April 10, 2019).
- 36 —, and Coauthors, 2018b: Risk of increased food insecurity under stringent global climate change
37 mitigation policy. *Nat. Clim. Chang.*, **8**, 699–703, doi:10.1038/s41558-018-0230-x.
- 38 Hashemi, S.M. and de Montesquiou, A., 2011: Reaching the Poorest: Lessons from the Graduation
39 Model. *Focus Note 69*.
- 40 Häyhä, T., and P. P. Franzese, 2014: Ecosystem services assessment: A review under an ecological-
41 economic and systems perspective. *Ecol. Modell.*, **289**, 124–132.
- 42 Head, B. W., 2014: Evidence, uncertainty, and wicked problems in climate change decision making in
43 Australia. *Environ. Plan. C Gov. Policy*, **32**, 663–679, doi:10.1068/c1240.
- 44 Headey, D., 2011: Rethinking the global food crisis: The role of trade shocks. *Food Policy*,
45 doi:10.1016/j.foodpol.2010.10.003.
- 46 —, J. Hoddinott, and S. Park, 2017: Accounting for nutritional changes in six success stories: A
47 regression-decomposition approach. *Glob. Food Sec.*, **13**, 12–20, doi:10.1016/j.gfs.2017.02.003.

- 1 Heck, V., D. Gerten, W. Lucht, and A. Popp, 2018a: Biomass-based negative emissions difficult to
2 reconcile with planetary boundaries. *Nat. Clim. Chang.*, **8**, 151–155, doi:10.1038/s41558-017-
3 0064-y. <http://www.nature.com/articles/s41558-017-0064-y> (Accessed April 10, 2019).
- 4 ———, ———, ———, and ———, 2018b: Biomass-based negative emissions difficult to reconcile with
5 planetary boundaries. *Nat. Clim. Chang.*, **8**, 151–155, doi:10.1038/s41558-017-0064-y.
- 6 HEI/IHME, 2018: *A Special Report on Global Exposure To Air Pollution and Its Disease Burden*.
7 Boston, 24 pp. http://www.stateofglobalair.org/sites/default/files/soga_2019_report.pdf
8 (Accessed April 14, 2019).
- 9 Hejazi, M. I., and Coauthors, 2014: Integrated assessment of global water scarcity over the 21st
10 century under multiple climate change mitigation policies. *Hydrol. Earth Syst. Sci.*, **18**, 2859–
11 2883, doi:10.5194/hess-18-2859-2014. <https://www.hydrol-earth-syst-sci.net/18/2859/2014/>
12 (Accessed April 10, 2019).
- 13 Helmke, G., and S. Levitsky, 2004: Informal Institutions and Comparative Politics: A Research
14 Agenda. *Perspect. Polit.*, **2**, 725–740, doi:10.1017/S1537592704040472.
- 15 Hember, R. A., W. A. Kurz, and N. C. Coops, 2017: Relationships between individual-tree mortality
16 and water-balance variables indicate positive trends in water stress-induced tree mortality across
17 North America. *Glob. Chang. Biol.*, **23**, 1691–1710, doi:10.1111/gcb.13428.
18 <http://doi.wiley.com/10.1111/gcb.13428> (Accessed April 14, 2019).
- 19 Henderson, B., 2018: *A GLOBAL ECONOMIC EVALUATION OF GHG MITIGATION POLICIES*
20 *FOR AGRICULTURE*.
- 21 Hendrix, C. S., and I. Salehyan, 2012: Climate change, rainfall, and social conflict in Africa. *J. Peace*
22 *Res.*, **49**, 35–50, doi:10.1177/0022343311426165.
- 23 Henriksen, H. J., M. J. Roberts, P. van der Keur, A. Harjanne, D. Egilson, and L. Alfonso, 2018:
24 Participatory early warning and monitoring systems: A Nordic framework for web-based flood
25 risk management. *Int. J. Disaster Risk Reduct.*, doi:10.1016/j.ijdr.2018.01.038.
- 26 Henstra, D., 2016: The tools of climate adaptation policy: analysing instruments and instrument
27 selection. *Clim. Policy*, **16**, 496–521, doi:10.1080/14693062.2015.1015946.
28 <https://doi.org/10.1080/14693062.2015.1015946>.
- 29 Herendeen, N., and N. Glazier, 2009: Agricultural best management practices for Conesus Lake: The
30 role of extension and soil/water conservation districts. *J. Great Lakes Res.*, **35**, 15–22,
31 doi:10.1016/j.jglr.2008.08.005.
- 32 Herman, J. D., H. B. Zeff, P. M. Reed, and G. W. Characklis, 2014: Beyond optimality:
33 Multistakeholder robustness tradeoffs for regional water portfolio planning under deep
34 uncertainty. *Water Resour. Res.*, **50**, 7692–7713, doi:10.1002/2014WR015338.
- 35 Hermann, A., Koflerl, P., Mairhofer, J. P., 2016: *Climate Risk Insurance: New Approaches and*
36 *Schemes*.
37 https://www.allianz.com/v_1479816006000/media/economic_research/publications/working_papers/en/ClimateRisk.pdf
38 (Accessed May 25, 2018).
- 39 Hernandez, R. R., M. K. Hoffacker, M. L. Murphy-Mariscal, G. C. Wu, and M. F. Allen, 2015: Solar
40 energy development impacts on land cover change and protected areas. *Proc. Natl. Acad. Sci.*,
41 **112**, 13579–13584.
- 42 Hertel, T. W., M. B. Burke, and D. B. Lobell, 2010: The poverty implications of climate-induced crop
43 yield changes by 2030. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2010.07.001.
- 44 Hewitt, K., I. R. Jahn, L. Guadagno, E. Gandia, V. Bonnefoy, and K. Hewitt, 2017: Identifying
45 Emerging Issues in Disaster Risk Reduction, Migration, Climate Change and Sustainable
46 Development. *Identifying Emerg. Issues Disaster Risk Reduction, Migr. Clim. Chang. Sustain.*
47 *Dev.*, doi:10.1007/978-3-319-33880-4.

- 1 Hewitt, R., H. van Delden, and F. Escobar, 2014: Participatory land use modelling, pathways to an
2 integrated approach. *Environ. Model. Softw.*, **52**, 149–165,
3 doi:10.1016/J.ENVSOF.2013.10.019.
4 <http://www.sciencedirect.com/science/article/pii/S1364815213002624> (Accessed December 12,
5 2017).
- 6 Higgins, S. A., I. Overeem, K. G. Rogers, and E. A. Kalina, 2018: River linking in India: Downstream
7 impacts on water discharge and suspended sediment transport to deltas. *Elem Sci Anth*, **6**, 20,
8 doi:10.1525/elementa.269. <http://www.elementascience.org/article/10.1525/elementa.269/>
9 (Accessed May 18, 2018).
- 10 Himanen, S. J., P. Rikkonen, and H. Kahiluoto, 2016: Codesigning a resilient food system. *Ecol. Soc.*,
11 **21**, doi:10.5751/ES-08878-210441.
- 12 Hinkel, J., and Coauthors, 2014: Coastal flood damage and adaptation costs under 21st century sea-
13 level rise. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1222469111.
- 14 Hirsch, A. L., and Coauthors, 2018: Biogeophysical Impacts of Land-Use Change on Climate
15 Extremes in Low-Emission Scenarios: Results From HAPPI-Land. *Earth's Futur.*, **6**,
16 doi:10.1002/2017EF000744.
- 17 Hjort, J., Karjalainen, O., Aalto, J., Westermann, S., Romanovsky, V. E., Nelson, F. E., ... Luoto, M.
18 (2018). Degrading permafrost puts Arctic infrastructure at risk by mid-century. *Nature*
19 *Communications*, 9(1). <https://doi.org/10.1038/s41467-018-07557-4>
- 20 Hoegh-Guldberg, O., and Coauthors, 2018: Impacts of 1.5°C Global Warming on Natural and Human
21 Systems. *Global Warming of 1.5°C. An IPCC Special Report on the impacts of global warming*
22 *of 1.5°C above pre-industrial levels and related global greenhouse gas emission pathways, in*
23 *the context of strengthening the global response to the threat of climate change.*
- 24 Hoeninghaus, D. J., A. A. Agostinho, L. C. Gomes, F. M. Pelicice, E. K. Okada, J. D. Latini, E. A. L.
25 Kashiwaqui, and K. O. Winemiller, 2009: Effects of river impoundment on ecosystem services
26 of large tropical rivers: Embodied energy and market value of artisanal fisheries. *Conserv. Biol.*,
27 **23**, 1222–1231, doi:10.1111/j.1523-1739.2009.01248.x.
- 28 Hoff, H., 2011: Understanding the Nexus. *Backgr. Pap. Bonn2011 Conf. Water, Energy Food Secur.*
29 *Nexus.*, 1–52.
- 30 van der Hoff, R., R. Rajão, and P. Leroy, 2018: Clashing interpretations of REDD+ “results” in the
31 Amazon Fund. *Clim. Change*, 1–13, doi:10.1007/s10584-018-2288-x.
32 <http://link.springer.com/10.1007/s10584-018-2288-x> (Accessed October 15, 2018).
- 33 Hoffmann, H., G. Uckert, C. Reif, K. Müller, and S. Sieber, 2015a: Traditional biomass energy
34 consumption and the potential introduction of firewood efficient stoves: insights from western
35 Tanzania. *Reg. Environ. Chang.*, **15**, 1191–1201, doi:10.1007/s10113-014-0738-1.
36 <http://link.springer.com/10.1007/s10113-014-0738-1> (Accessed April 14, 2019).
- 37 —, —, —, —, and —, 2015b: Traditional biomass energy consumption and the potential
38 introduction of firewood efficient stoves: insights from western Tanzania. *Reg. Environ. Chang.*,
39 **15**, 1191–1201, doi:10.1007/s10113-014-0738-1.
- 40 Hoffmann, H. K., K. Sander, M. Brüntrup, and S. Sieber, 2017: Applying the Water-Energy-Food
41 Nexus to the Charcoal Value Chain. *Front. Environ. Sci.*, **5**, 84, doi:10.3389/fenvs.2017.00084.
42 <http://journal.frontiersin.org/article/10.3389/fenvs.2017.00084/full> (Accessed April 14, 2019).
- 43 Höhne, N., and Coauthors, 2017: The Paris Agreement: resolving the inconsistency between global
44 goals and national contributions. *Clim. Policy*, **17**, 16–32, doi:10.1080/14693062.2016.1218320.
- 45 Hojas-Gascon, L., H. D. Eva, D. Ehrlich, M. Pesaresi, F. Achard, and J. Garcia, 2016a: Urbanization
46 and forest degradation in east Africa - a case study around Dar es Salaam, Tanzania. *2016 IEEE*
47 *International Geoscience and Remote Sensing Symposium (IGARSS)*, IEEE, 7293–7295
48 <http://ieeexplore.ieee.org/document/7730902/> (Accessed April 14, 2019).

- 1 ———, ———, ———, ———, ———, and ———, 2016b: Urbanization and forest degradation in east Africa -
2 a case study around Dar es Salaam, Tanzania. *2016 IEEE International Geoscience and Remote*
3 *Sensing Symposium (IGARSS)*, IEEE, 7293–7295.
- 4 Holden, E., K. Linnerud, and D. Banister, 2017: The Imperatives of Sustainable Development.
5 *Sustain. Dev.*, **25**, 213–226, doi:10.1002/sd.1647. <http://doi.wiley.com/10.1002/sd.1647>.
- 6 Holden, S. T., and H. Ghebru, 2016: Land tenure reforms, tenure security and food security in poor
7 agrarian economies: Causal linkages and research gaps. *Glob. Food Sec.*, **10**, 21–28,
8 doi:10.1016/j.gfs.2016.07.002.
- 9 Holland, M. B., K. W. Jones, L. Naughton-Treves, J. L. Freire, M. Morales, and L. Suárez, 2017:
10 Titling land to conserve forests: The case of Cuyabeno Reserve in Ecuador. *Glob. Environ.*
11 *Chang.*, **44**, 27–38, doi:10.1016/j.gloenvcha.2017.02.004.
- 12 Holling, C. S., 1978: Adaptive Environmental Assessment and Management. *Int. Ser. Appl. Syst.*
13 *Anal.*, 402.
- 14 Holling, C. S., 1986: Adaptive environmental management. *Environment*, **28**, 39,
15 doi:10.1080/00139157.1986.9928829.
- 16 Holstenkamp, L., and F. Kahla, 2016: What are community energy companies trying to accomplish?
17 An empirical investigation of investment motives in the German case. *Energy Policy*, **97**, 112–
18 122, doi:10.1016/j.enpol.2016.07.010.
- 19 Hordijk, M., L. M. Sara, and C. Sutherland, 2014: Resilience, transition or transformation? A
20 comparative analysis of changing water governance systems in four southern cities. *Environ.*
21 *Urban.*, doi:10.1177/0956247813519044.
- 22 Hornsey, M. J., E. A. Harris, P. G. Bain, and K. S. Fielding, 2016: Meta-analyses of the determinants
23 and outcomes of belief in climate change. *Nat. Clim. Chang.*, **6**, 622–626,
24 doi:10.1038/nclimate2943.
- 25 Hossain, M., 2018: Introduction: Pathways to a Sustainable Economy. *Pathways to a Sustainable*
26 *Economy*, Springer International Publishing, Cham, 1–11 [http://link.springer.com/10.1007/978-](http://link.springer.com/10.1007/978-3-319-67702-6_1)
27 [3-319-67702-6_1](http://link.springer.com/10.1007/978-3-319-67702-6_1) (Accessed April 3, 2019).
- 28 Hou, D., 2016: Divergence in stakeholder perception of sustainable remediation. *Sustain. Sci.*, **11**,
29 215–230, doi:10.1007/s11625-015-0346-0.
- 30 ———, and A. Al-Tabbaa, 2014: Sustainability: A new imperative in contaminated land remediation.
31 *Environ. Sci. Policy*, **39**, 25–34, doi:10.1016/j.envsci.2014.02.003.
- 32 Howlett, M., and J. Rayner, 2013: Patching vs Packaging in Policy Formulation: Assessing Policy
33 Portfolio Design. *Polit. Gov.*, **1**, 170, doi:10.17645/pag.v1i2.95.
- 34 Huang, J., and G. Yang, 2017: Understanding recent challenges and new food policy in China. *Glob.*
35 *Food Sec.*, doi:10.1016/j.gfs.2016.10.002.
- 36 Huang, J., H. Yu, X. Guan, G. Wang, and R. Guo, 2016: Accelerated dryland expansion under
37 climate change. *Nat. Clim. Chang.*, **6**, 166–171, doi:10.1038/nclimate2837.
38 <http://www.nature.com/articles/nclimate2837> (Accessed April 9, 2019).
- 39 ———, ———, A. Dai, Y. Wei, and L. Kang, 2017: Drylands face potential threat under 2 °C global
40 warming target. *Nat. Clim. Chang.*, **7**, 417–422, doi:10.1038/nclimate3275.
41 <http://www.nature.com/articles/nclimate3275> (Accessed April 11, 2019).
- 42 Huang, R., D. Tian, J. Liu, S. Lv, X. He, and M. Gao, 2018: Responses of soil carbon pool and soil
43 aggregates associated organic carbon to straw and straw-derived biochar addition in a dryland
44 cropping mesocosm system. *Agric. Ecosyst. Environ.*, **265**, 576–586,
45 doi:10.1016/J.AGEE.2018.07.013. [https://www.sciencedirect-](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0167880918302901?via%3Dihub)
46 [com.ezp.sub.su.se/science/article/pii/S0167880918302901?via%3Dihub](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0167880918302901?via%3Dihub) (Accessed April 14,
47 2019).

- 1 Huber-Sannwald, E., M. R. Palacios, J. T. A. Moreno, M. Braasch, R. M. M. Peña, J. G. de A.
2 Verduzco, and K. M. Santos, 2012: Navigating challenges and opportunities of land degradation
3 and sustainable livelihood development in dryland social-ecological systems: A case study from
4 Mexico. *Philos. Trans. R. Soc. B Biol. Sci.*, doi:10.1098/rstb.2011.0349.
- 5 Hudiburg, T. W., S. C. Davis, W. Parton, and E. H. Delucia, 2015: Bioenergy crop greenhouse gas
6 mitigation potential under a range of management practices. *GCB Bioenergy*,
7 doi:10.1111/gcbb.12152.
- 8 Hudson, P., W. J. W. Botzen, L. Feyen, and J. C. J. H. Aerts, 2016: Incentivising flood risk adaptation
9 through risk based insurance premiums: Trade-offs between affordability and risk reduction.
10 *Ecol. Econ.*, **125**, 1–13, doi:10.1016/J.ECOLECON.2016.01.015.
11 <https://www.sciencedirect.com/science/article/pii/S0921800916301240> (Accessed May 16,
12 2018).
- 13 Hugo, G. J., 2011: Lessons from Past Forced Resettlement for Climate Change Migration. *Migration
14 and Climate Change*.
- 15 Huisheng, S., 2015: Between the Formal and Informal: Institutions and Village Governance in Rural
16 China. *An Int. J.*, **13**, 24–44. <https://muse.jhu.edu/article/589970> (Accessed December 22, 2017).
- 17 Humpenöder, F., and Coauthors, 2017: Large-scale bioenergy production: How to resolve
18 sustainability trade-offs? *Environ. Res. Lett.*, doi:10.1088/1748-9326/aa9e3b.
19 <http://iopscience.iop.org/article/10.1088/1748-9326/aa9e3b> (Accessed December 8, 2017).
- 20 Hunsberger, C., and Coauthors, 2017: Climate change mitigation, land grabbing and conflict: towards
21 a landscape-based and collaborative action research agenda. *Can. J. Dev. Stud.*, **38**, 305–324,
22 doi:10.1080/02255189.2016.1250617.
- 23 Hunzai, K., T. Chagas, L. Gilde, T. Hunzai, and N. Krämer, 2018: *Finance options and instruments
24 for Ecosystem-based Adaptation. Overview and compilation of ten examples*. Bonn,.
- 25 Hurlbert, M., 2015a: Climate justice: A call for leadership. *Environ. Justice*, **8**, 51–55,
26 doi:10.1089/env.2014.0035.
- 27 —, 2015b: Learning, participation, and adaptation: exploring agri-environmental programmes. *J.
28 Environ. Plan. Manag.*, **58**, 113–134, doi:10.1080/09640568.2013.847823.
- 29 —, 2018a: The challenge of integrated flood risk governance: case studies in Alberta and
30 Saskatchewan, Canada. *Int. J. River Basin Manag.*, doi:10.1080/15715124.2018.1439495.
- 31 —, and J. Pittman, 2014: Exploring adaptive management in environmental farm programs in
32 Saskatchewan, Canada. *J. Nat. Resour. Policy Res.*, **6**, 195–212,
33 doi:10.1080/19390459.2014.915131.
- 34 —, and J. Gupta, 2015: The split ladder of participation: A diagnostic, strategic, and evaluation tool
35 to assess when participation is necessary. *Environ. Sci. Policy*, **50**, 100–113,
36 doi:10.1016/j.envsci.2015.01.011.
- 37 —, and —, 2016: Adaptive Governance, Uncertainty, and Risk: Policy Framing and Responses
38 to Climate Change, Drought, and Flood. *Risk Anal.*, **36**, 339–356, doi:10.1111/risa.12510.
- 39 —, and P. Mussetta, 2016: Creating resilient water governance for irrigated producers in Mendoza,
40 Argentina. *Environ. Sci. Policy*, **58**, 83–94, doi:10.1016/j.envsci.2016.01.004.
- 41 —, and J. Gupta, 2017: The adaptive capacity of institutions in Canada, Argentina, and Chile to
42 droughts and floods. *Reg. Environ. Chang.*, **17**, doi:10.1007/s10113-016-1078-0.
- 43 Hurlbert, M. A., 2018b: *Adaptive Governance of Disaster: Drought and Flood in Rural Areas*.
44 Springer, Cham, Switzerland,.
- 45 Hurlbert, M. A., and H. Diaz, 2013: Water governance in Chile and Canada: A comparison of
46 adaptive characteristics. *Ecol. Soc.*, **18**, doi:10.5751/ES-06148-180461.

- 1 ———, and E. Montana, 2015: Dimensions of adaptive water governance and drought in Argentina and
2 Canada. *J. Sustain. Dev.*, **8**, 120–137, doi:10.5539/jsd.v8n1p120.
- 3 Hurlbert, M. A., and J. Gupta, 2018: An institutional analysis method for identifying policy
4 instruments facilitating the adaptive governance of drought. *Environ. Sci. Policy*, **forthcomin**.
- 5 Hurlimann, A. C., and A. P. March, 2012: The role of spatial planning in adapting to climate change.
6 *Wiley Interdiscip. Rev. Clim. Chang.*, **3**, 477–488, doi:10.1002/wcc.183.
- 7 Huttunen, S., P. Kivimaa, and V. Virkamäki, 2014: The need for policy coherence to trigger a
8 transition to biogas production. *Environ. Innov. Soc. Transitions*, **12**, 14–30,
9 doi:10.1016/j.eist.2014.04.002.
- 10 Huyer, S., J. Twyman, M. Koningstein, J. Ashby, and S. Vermeulen, 2015a: Supporting women
11 farmers in a changing climate: five policy lessons. *CGIAR, CCAFS*,
12 [https://ccafs.cgiar.org/publications/supporting-women-farmers-changing-climate-five-policy-](https://ccafs.cgiar.org/publications/supporting-women-farmers-changing-climate-five-policy-lessons#.XKNBVJgzY2w)
13 [lessons#.XKNBVJgzY2w](https://ccafs.cgiar.org/publications/supporting-women-farmers-changing-climate-five-policy-lessons#.XKNBVJgzY2w) (Accessed April 2, 2019).
- 14 ———, ———, ———, ———, and ———, 2015b: Supporting women farmers in a changing climate: five
15 policy lessons. *CGIAR, CCAFS*,.
- 16 Iacobuta, G., N. K. Dubash, P. Upadhyaya, M. Deribe, and N. Höhne, 2018: National climate change
17 mitigation legislation, strategy and targets: a global update. *Clim. Policy*,
18 doi:10.1080/14693062.2018.1489772.
- 19 ICCC, 2018: *Interim Climate Change Committee Terms of Reference and Appointment*. New Zealand,
20 [http://www.mfe.govt.nz/sites/default/files/media/Legislation/Cabinet](http://www.mfe.govt.nz/sites/default/files/media/Legislation/Cabinet_paper/interim-climate-change-committee-tor.pdf) paper/interim-climate-
21 [change-committee-tor.pdf](http://www.mfe.govt.nz/sites/default/files/media/Legislation/Cabinet_paper/interim-climate-change-committee-tor.pdf).
- 22 ICSU, 2017: *A Guide to SDG Interactions: from Science to Implementation*. Paris,.
- 23 IEA, 2017: *World Energy Outlook 2017*. International Energy Agency, Paris,.
- 24 Ighodaro, I. D., F. S. Lategan, and W. Mupindu, 2016: The Impact of Soil Erosion as a Food Security
25 and Rural Livelihoods Risk in South Africa. *J. Agric. Sci.*, **8**, 1, doi:10.5539/jas.v8n8p1.
- 26 Iglesias, A., and L. Garrote, 2015: Adaptation strategies for agricultural water management under
27 climate change in Europe. *Agric. Water Manag.*, doi:10.1016/j.agwat.2015.03.014.
- 28 Iizumi, T., H. Sakuma, M. Yokozawa, J. J. Luo, A. J. Challinor, M. E. Brown, G. Sakurai, and T.
29 Yamagata, 2013: Prediction of seasonal climate-induced variations in global food production.
30 *Nat. Clim. Chang.*, doi:10.1038/nclimate1945.
- 31 Innocenti, D., and P. Albrito, 2011: Reducing the risks posed by natural hazards and climate change:
32 The need for a participatory dialogue between the scientific community and policy makers.
33 *Environ. Sci. Policy*, doi:10.1016/j.envsci.2010.12.010.
- 34 Internal Displacement Monitoring Center, 2017: GLOBAL DISASTER DISPLACEMENT RISK - A
35 BASELINE FOR FUTURE WORK. *IDMC*,.
- 36 IPBES, 2018: *The assessment report on Land Degradation and Restoration*.
37 [https://www.ipbes.net/system/tdf/2018_ldr_full_report_book_v4_pages.pdf?file=1&type=node](https://www.ipbes.net/system/tdf/2018_ldr_full_report_book_v4_pages.pdf?file=1&type=node&id=29395)
38 [&id=29395](https://www.ipbes.net/system/tdf/2018_ldr_full_report_book_v4_pages.pdf?file=1&type=node&id=29395).
- 39 IPCC, 2001: CLIMATE CHANGE 2001: impacts, adaptation and vulnerability. *Work. Gr. II*
40 *Impacts Adapt. vulnerability*, 10, doi:10.1002/joc.775.
- 41 ———, 2012: *Managing the Risks of Extreme Events and Disasters to Advance Climate Change*
42 *Adaptation. A Special Report of Working Groups I and II of the Intergovernmental Panel on*
43 *Climate Change*. Cambridge and New York, [https://www.ipcc.ch/pdf/special-](https://www.ipcc.ch/pdf/special-reports/srex/SREX_Full_Report.pdf)
44 [reports/srex/SREX_Full_Report.pdf](https://www.ipcc.ch/pdf/special-reports/srex/SREX_Full_Report.pdf) (Accessed October 31, 2018).
- 45 ———, 2014a: *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and*
46 *Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the*

- 1 *Intergovernmental Panel on Climate Change*. C.B. [Field et al., Eds. Cambridge University
2 Press, Cambridge,.
- 3 —, 2014b: Summary for policymakers. In: *Climate Change 2014: Impacts, Adaptation, and*
4 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the*
5 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. And L.L.W. Field,
6 C.B., V.R. Barros, D.J. Dokken, K.J. Mach, M.D. Mastrandrea, T.E. Bilir, M. Chatterjee, K.L.
7 Ebi, Y.O. Estrada, R.C. Genova, B. Girma, E.S. Kissel, A.N. Levy, S. MacCracken, P.R.
8 Mastrandrea, and L.L. White Field, C.B., V.R. Barros, D.J. Dokken, K., Ed., Cambridge
9 University Press, Cambridge, United Kingdom and New York, NY, USA.
- 10 —, 2014c: *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to*
11 *the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. 151 pp.
- 12 —, 2018a: Summary for Policy Makers. In: *Global Warming of 1.5°C*.
13 <http://www.ipcc.ch/report/sr15/>.
- 14 —, 2018b: Summary for Policy Makers. *IPCC Special Report on the impacts of global warming of*
15 *1.5°C*, William Solecki http://report.ipcc.ch/sr15/pdf/sr15_spm_final.pdf (Accessed November
16 1, 2018).
- 17 —, 2018c: Summary for Policymakers. *Global Warming of 1.5 °C an IPCC special report on the*
18 *impacts of global warming of 1.5 °C above pre-industrial levels and related global greenhouse*
19 *gas emission pathways, in the context of strengthening the global response to the threat of*
20 *climate change* http://report.ipcc.ch/sr15/pdf/sr15_spm_final.pdf (Accessed October 29, 2018).
- 21 —, 2018d: *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of*
22 *1.5 °C above pre-industrial levels and related global greenhouse gas emission pathways, in the*
23 *context of strengthening the global response to the threat of climate change*.
24 <http://www.ipcc.ch/report/sr15/>.
- 25 —, 2018e: Summary for Policymakers. *Global Warming of 1.5°C. An IPCC Special Report on the*
26 *impacts of global warming of 1.5°C above pre-industrial levels and related global greenhouse*
27 *gas emission pathways, in the context of strengthening the global response to the threat of*
28 *climate change*, H.-O. Masson-Delmotte, V., P. Zhai, D. Pörtner, A.P. Roberts, J. Skea, P.R.
29 Shukla, X. W. Moufouma-Okia, C. Péan, R. Pidcock, S. Connors, J.B.R. Matthews, Y. Chen,
30 M.I. Zhou, E.L. Gomis, and And T.W. T. Maycock, M. Tignor, Eds., World Meteorological
31 Organization, Geneva, Switzerland, p. 32.
- 32 IPCC 2014, 2014: *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and*
33 *III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*. IPCC,
34 Ed. Gian-Kasper Plattner, Geneva, <http://www.ipcc.ch>. (Accessed October 31, 2018).
- 35 ISEAL Alliance, Private sustainability standards and the EU Renewable Energy Directive. 2018.,
36 [https://www.isealalliance.org/impacts-and-benefits/case-studies/private-sustainability-standards-](https://www.isealalliance.org/impacts-and-benefits/case-studies/private-sustainability-standards-and-eu-renewable-energy)
37 [and-eu-renewable-energy](https://www.isealalliance.org/impacts-and-benefits/case-studies/private-sustainability-standards-and-eu-renewable-energy) (Accessed April 8, 2019a).
- 38 —, Private sustainability standards and the EU Renewable Energy Directive. 2018.,
- 39 Ishida, H., and Coauthors, 2014: Global-scale projection and its sensitivity analysis of the health
40 burden attributable to childhood undernutrition under the latest scenario framework for climate
41 change research. *Environ. Res. Lett.*, **9**, 064014, doi:10.1088/1748-9326/9/6/064014.
42 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/9/i=6/a=064014?key=crossref.288e05e1c145e8fa5c1873d6d19c9811)
43 [9326/9/i=6/a=064014?key=crossref.288e05e1c145e8fa5c1873d6d19c9811](http://stacks.iop.org/1748-9326/9/i=6/a=064014?key=crossref.288e05e1c145e8fa5c1873d6d19c9811) (Accessed April 14,
44 2019).
- 45 Ismail, F., F. Amir, A. Md, H. S. Jaffer, K. Alajaji, L. Unwin, Y. Yee, and T. S. Jamil, 2017: *Market*
46 *trends in family and general Takaful*.
47 <http://www.milliman.com/uploadedFiles/insight/2017/Takaful-2017-full-report.pdf> (Accessed
48 April 3, 2019).
- 49 ISO, 2009: *Australia and New Zealand Risk Management Standards 31000:2009*. AS/NZS,.

- 1 —, 2015: *ISO 13065:2015 - Sustainability criteria for bioenergy*. 57 pp.
2 <https://www.iso.org/standard/52528.html> (Accessed April 8, 2019).
- 3 —, 2017: *Environmental management -- Guidelines for establishing good practices for combatting*
4 *land degradation and desertification -- Part 1: Good practices framework*. 31 pp.
5 <https://www.iso.org/standard/64646.html>.
- 6 IWAI, 2016: *Consolidated Environmental Impact Assessment Report of National Waterways-1 :*
7 *Volume – 3*. [http://documents.worldbank.org/curated/en/190981468255890798/pdf/SFG2231-](http://documents.worldbank.org/curated/en/190981468255890798/pdf/SFG2231-V1-REVISED-EA-P148775-Box396336B-PUBLIC-Disclosed-12-5-2016.pdf)
8 [V1-REVISED-EA-P148775-Box396336B-PUBLIC-Disclosed-12-5-2016.pdf](http://documents.worldbank.org/curated/en/190981468255890798/pdf/SFG2231-V1-REVISED-EA-P148775-Box396336B-PUBLIC-Disclosed-12-5-2016.pdf).
- 9 Iyahen, E., and J. Syroka, 2018: Managing Risks from Climate Change on the African Continent: The
10 African Risk Capacity (ARC) as an Innovative Risk Financing Mechanism. *Resilience: The*
11 *Science of Adaptation to Climate Change*, Z. Zommers and K. Alverson, Eds., Elsevier.
- 12 Jagger, P., and J. Pender, 2006: Influences of programs and organizations on the adoption of
13 sustainable land management technologies in Uganda. *Strategies for sustainable land*
14 *management in the East African highlands*, J. Pender, F. Place, S. Ehui, and I.F.P.R. Institute,
15 Eds., International Food Policy Research Institute., Washington, D.C, 277–306.
- 16 Jagger, P., and N. Kittner, 2017: Deforestation and biomass fuel dynamics in Uganda. *Biomass and*
17 *Bioenergy*, **105**, 1–9, doi:10.1016/j.biombioe.2017.06.005.
18 <http://dx.doi.org/10.1016/j.biombioe.2017.06.005> (Accessed April 15, 2019).
- 19 —, and I. Das, 2018: Implementation and scale-up of a biomass pellet and improved cookstove
20 enterprise in Rwanda. *Energy Sustain. Dev.*, **46**, 32–41, doi:10.1016/j.esd.2018.06.005.
21 <https://doi.org/10.1016/j.esd.2018.06.005> (Accessed April 15, 2019).
- 22 Jalbert, K., and A. J. Kinchy, 2016: Sense and Influence: Environmental Monitoring Tools and the
23 Power of Citizen Science. *J. Environ. Policy Plan.*, **18**, 379–397,
24 doi:10.1080/1523908X.2015.1100985.
25 <http://www.tandfonline.com/doi/full/10.1080/1523908X.2015.1100985> (Accessed December 11,
26 2017).
- 27 Jaleta, M., M. Kassie, and B. Shiferaw, 2013: Tradeoffs in crop residue utilization in mixed crop-
28 livestock systems and implications for conservation agriculture. *Agric. Syst.*,
29 doi:10.1016/j.agsy.2013.05.006.
- 30 —, —, and O. Erenstein, 2015: Determinants of maize stover utilization as feed, fuel and soil
31 amendment in mixed crop-livestock systems, Ethiopia. *Agric. Syst.*,
32 doi:10.1016/j.agsy.2014.08.010.
- 33 James, J., R. Harrison, J. James, and R. Harrison, 2016: The Effect of Harvest on Forest Soil Carbon:
34 A Meta-Analysis. *Forests*, **7**, 308, doi:10.3390/f7120308. [http://www.mdpi.com/1999-](http://www.mdpi.com/1999-4907/7/12/308)
35 [4907/7/12/308](http://www.mdpi.com/1999-4907/7/12/308) (Accessed April 14, 2019).
- 36 James, R., R. Washington, C. F. Schleussner, J. Rogelj, and D. Conway, 2017: Characterizing half-a-
37 degree difference: a review of methods for identifying regional climate responses to global
38 warming targets. *Wiley Interdiscip. Rev. Clim. Chang.*, doi:10.1002/wcc.457.
- 39 Janetos, A., C. Justice, M. Jahn, M. Obersteiner, J. Glauber, and W. Mulhern, 2017: *The Risks of*
40 *Multiple Breadbasket Failures in the 21st Century: A Science Research Agenda*. 24 pp.
- 41 Janif, S. Z., P. D. Nunn, P. Geraghty, W. Aalbersberg, F. R. Thomas, and M. Camailakeba, 2016:
42 Value of traditional oral narratives in building climate-change resilience: insights from rural
43 communities in Fiji. *Ecol. Soc.*, **21**, art7, doi:10.5751/ES-08100-210207.
44 <http://www.ecologyandsociety.org/vol21/iss2/art7/> (Accessed April 7, 2019).
- 45 Jansujwicz, J. S., A. J. K. Calhoun, and R. J. Lillieholm, 2013: The maine vernal pool mapping and
46 assessment program: Engaging municipal officials and private landowners in community-based
47 citizen science. *Environ. Manage.*, **52**, 1369–1385, doi:10.1007/s00267-013-0168-8.
- 48 Jantarasami, L. C., J. J. Lawler, and C. W. Thomas, 2010: Institutional barriers to climate change

- 1 adaptation in U.S. National parks and forests. *Ecol. Soc.*, doi:10.5751/ES-03715-150433.
- 2 Jeffrey, S. R., D. E. Trautman, and J. R. Unterschultz, 2017: Canadian Agricultural Business Risk
3 Management Programs: Implications for Farm Wealth and Environmental Stewardship. *Can. J.*
4 *Agric. Econ. Can. d'agroéconomie*, **65**, 543–565, doi:10.1111/cjag.12145.
5 <http://doi.wiley.com/10.1111/cjag.12145>.
- 6 Jelsma, I., M. Slingerland, K. Giller, J. B.-J. of R. Studies, and undefined 2017, Collective action in a
7 smallholder oil palm production system in Indonesia: The key to sustainable and inclusive
8 smallholder palm oil? *Elsevier*,
9 <https://www.sciencedirect.com/science/article/pii/S0743016716303941> (Accessed May 23,
10 2018).
- 11 Jepson, E. J., and A. L. Haines, 2014: Zoning for Sustainability: A Review and Analysis of the
12 Zoning Ordinances of 32 Cities in the United States. *J. Am. Plan. Assoc.*, **80**, 239–252,
13 doi:10.1080/01944363.2014.981200.
14 <http://www.tandfonline.com/doi/abs/10.1080/01944363.2014.981200> (Accessed February 25,
15 2019).
- 16 Jeuland, M., S. K. Pattanayak, and R. Bluffstone, 2015: The Economics of Household Air Pollution.
17 *Annu. Rev. Resour. Econ.*, **7**, 81–108, doi:10.1146/annurev-resource-100814-125048.
18 <http://www.annualreviews.org/doi/10.1146/annurev-resource-100814-125048> (Accessed April
19 15, 2019).
- 20 Jiang, J., W. Wang, C. Wang, and Y. Liu, 2017: Chinese Journal of Population Resources and
21 Environment Combating climate change calls for a global technological cooperation system
22 built on the concept of ecological civilization) Combating climate change calls for a global
23 technological cooperation system built on the concept of ecological civilization Combating
24 climate change calls for a global technological cooperation system built on the concept of
25 ecological civilization. *Chinese J. Popul. Resour. Environ.*, **15**, 21–31,
26 doi:10.1080/10042857.2017.1286145.
27 <http://www.tandfonline.com/action/journalInformation?journalCode=tpre20> (Accessed May 22,
28 2018).
- 29 Jjemba, E. W., B. K. Mwebaze, J. Arrighi, E. Coughlan de Perez, and M. Bailey, 2018: Forecast-
30 Based Financing and Climate Change Adaptation: Uganda Makes History Using Science to
31 Prepare for Floods. *Resilience: The Science of Adaptation to Climate Change*, K. Alverson and
32 Z. Zommers, Eds., Elsevier, Oxford, 237–243.
- 33 Joana Specht, M., S. Rodrigo Ribeiro Pinto, U. Paulino Albuquerque, M. Tabarelli, F. P. Melo, and D.
34 Manuel de Medeiros, 2015a: Burning biodiversity: Fuelwood harvesting causes forest
35 degradation in human-dominated tropical landscapes. *Glob. Ecol. Conserv.*, **3**, 200–209,
36 doi:10.1016/j.gecco.2014.12.002. <http://dx.doi.org/10.1016/j.gecco.2014.12.002> (Accessed April
37 15, 2019).
- 38 —, —, —, —, —, and —, 2015b: Burning biodiversity: Fuelwood harvesting causes
39 forest degradation in human-dominated tropical landscapes. *Glob. Ecol. Conserv.*, **3**, 200–209,
40 doi:10.1016/j.gecco.2014.12.002.
- 41 Johnson, B. B., and M. L. Becker, 2015: Social-ecological resilience and adaptive capacity in a
42 transboundary ecosystem. *Soc. Nat. Resour.*, doi:10.1080/08941920.2015.1037035.
- 43 Johnson, F. A., 2011a: Learning and adaptation in the management of waterfowl harvests. *J. Environ.*
44 *Manage.*, **92**, 1385–1394, doi:10.1016/j.jenvman.2010.10.064.
- 45 Johnson, F. X., 2011b: Regional-global Linkages in the Energy-Climate-Development Policy Nexus:
46 The Case of Biofuels in the EU Renewable Energy Directive. *Renew. Energy Law Policy Rev.*,
47 **2**, 91–106, doi:10.2307/24324724. <http://www.jstor.org.ezp.sub.su.se/stable/24324724>
48 (Accessed May 23, 2018).
- 49 —, 2017: Biofuels, Bioenergy and the Bioeconomy in North and South. *Ind. Biotechnol.*, **13**, 289–

- 1 doi:10.1016/j.esd.2016.02.009. <http://dx.doi.org/10.1016/j.esd.2016.02.009> (Accessed April 15,
2 2019).
- 3 —, —, and —, 2016b: Charcoal as a diversification strategy: The flexible role of charcoal
4 production in the livelihoods of smallholders in central Mozambique.
5 doi:10.1016/j.esd.2016.02.009.
- 6 Jones, N., and E. Presler-Marshall, 2015: Cash Transfers. *International Encyclopedia of the Social &*
7 *Behavioral Sciences: Second Edition*.
- 8 Jones, R. N., and Coauthors, 2014: 95 2 Foundations for Decision Making Coordinating Lead
9 Authors: Lead Authors: Volunteer Chapter Scientist. 195–228.
10 http://www.ipcc.ch/pdf/assessment-report/ar5/wg2/WGIIAR5-Chap2_FINAL.pdf (Accessed
11 December 19, 2017).
- 12 Jongman, B., and Coauthors, 2014: Increasing stress on disaster-risk finance due to large floods. *Nat.*
13 *Clim. Chang.*, **4**, 264–268, doi:10.1038/nclimate2124.
- 14 Jordan, A. J., and Coauthors, 2015a: Emergence of polycentric climate governance and its future
15 prospects. *Nat. Clim. Chang.*, **5**, 977–982, doi:10.1038/nclimate2725.
16 <http://www.nature.com/doi/10.1038/nclimate2725> (Accessed December 21, 2017).
- 17 Jordan, R., A. Crall, S. Gray, T. Phillips, and D. Mellor, 2015b: Citizen science as a distinct field of
18 inquiry. *Bioscience*, **65**, 208–211, doi:10.1093/biosci/biu217.
- 19 Jost, C., and Coauthors, 2016: Understanding gender dimensions of agriculture and climate change in
20 smallholder farming communities. *Clim. Dev.*, **8**, 133–144,
21 doi:10.1080/17565529.2015.1050978.
- 22 Juhola, S., E. Glaas, B. O. Linnér, and T. S. Neset, 2016: Redefining maladaptation. *Environ. Sci.*
23 *Policy*, doi:10.1016/j.envsci.2015.09.014.
- 24 Jumani, S., S. Rao, S. Machado, and A. Prakash, 2017: Big concerns with small projects: Evaluating
25 the socio-ecological impacts of small hydropower projects in India. *Ambio*, **46**, 500–511,
26 doi:10.1007/s13280-016-0855-9. <http://www.ncbi.nlm.nih.gov/pubmed/28074405> (Accessed
27 May 18, 2018).
- 28 —, —, N. Kelkar, S. Machado, J. Krishnaswamy, and S. Vaidyanathan, 2018: Fish community
29 responses to stream flow alterations and habitat modifications by small hydropower projects in
30 the Western Ghats biodiversity hotspot, India. *Aquat. Conserv. Mar. Freshw. Ecosyst.*, **28**, 979–
31 993.
- 32 Junior, S., S. R. Santos, M. Travassos, and M. Vianna, 2012: Impact on a fish assemblage of the
33 maintenance dredging of a navigation channel in a tropical coastal ecosystem. *Brazilian J.*
34 *Oceanogr.*, **60**, 25–32.
- 35 Júnior, W. S. F., F. R. Santoro, I. Vandebroek, and U. P. Albuquerque, 2016: Urbanization,
36 Modernization, and Nature Knowledge. *Introduction to Ethnobiology*, Springer International
37 Publishing, Cham, 251–256 http://link.springer.com/10.1007/978-3-319-28155-1_37 (Accessed
38 April 9, 2019).
- 39 Jürisoo, M., F. Lambe, and M. Osborne, 2018: Beyond buying: The application of service design
40 methodology to understand adoption of clean cookstoves in Kenya and Zambia. *Energy Res.*
41 *Soc. Sci.*, **39**, 164–176, doi:10.1016/j.erss.2017.11.023.
42 <https://doi.org/10.1016/j.erss.2017.11.023> (Accessed April 15, 2019).
- 43 Kabeer, N., K. Mumtaz, and A. Sayeed, 2010: Beyond risk management: Vulnerability, social
44 protection and citizenship in Pakistan. *J. Int. Dev.*, **22**, 1–19, doi:10.1002/jid.1538.
- 45 Kaenzig, R., and E. Piguet, 2014: Migration and Climate Change in Latin America and the Caribbean.
46 *People on the Move in a Changing Climate. The Regional Impact of Environmental Change on*
47 *Migration*.

- 1 Kaijser, A., and A. Kronsell, 2014: Climate change through the lens of intersectionality. *Env. Polit.*,
2 **23**, 417–433, doi:10.1080/09644016.2013.835203.
3 <http://www.tandfonline.com/doi/abs/10.1080/09644016.2013.835203> (Accessed April 2, 2019).
- 4 Kainuma, M., K. Miwa, T. Ehara, O. Akashi, and Y. Asayama, 2013: A low-carbon society: Global
5 visions, pathways, and challenges. *Clim. Policy*, doi:10.1080/14693062.2012.738016.
- 6 Kaisa, K. K., B. Maria, M. Efrian, J. Sirkku, M. Moira, M. Cynthia, and D. Bimo, 2017: Analyzing
7 REDD+ as an experiment of transformative climate governance: Insights from Indonesia.
8 *Environ. Sci. Policy*, doi:10.1016/j.envsci.2017.03.014.
- 9 Kakota, T., D. Nyariki, D. Mkwambisi, and W. Kogi-Makau, 2011: Gender vulnerability to climate
10 variability and household food insecurity. *Clim. Dev.*, **3**, 298–309,
11 doi:10.1080/17565529.2011.627419.
- 12 Kale, E., 2017: Problematic uses and practices of farm ponds in Maharashtra. *Econ. Polit. Wkly.*, **52**,
13 20–22.
- 14 Kallbekken, S., and H. Sælen, 2013: ‘Nudging’ hotel guests to reduce food waste as a win–win
15 environmental measure. *Econ. Lett.*, **119**, 325–327,
16 doi:https://doi.org/10.1016/j.econlet.2013.03.019.
17 <http://www.sciencedirect.com/science/article/pii/S0165176513001286>.
- 18 Kangalawe.R.Y.M, Noe.C, Tungaraza.F.S.K, N. . & M. ., 2014: Understanding of Traditional
19 Knowledge and Indigenous Institutions on Sustainable Land Management in Kilimanjaro
20 Region , Tanzania. *Open J. Soil Sci.*, **4**, 469–493,
21 doi:http://dx.doi.org/10.4236/ojss.2014.413046.
- 22 Kanta Kafle, S., 2017: Disaster Early Warning Systems in Nepal: Institutional and Operational
23 Frameworks. *J. Geogr. Nat. Disasters*, doi:10.4172/2167-0587.1000196.
- 24 Kanter, D. R., and Coauthors, 2016: Evaluating agricultural trade-offs in the age of sustainable
25 development. *Agric. Syst.*, doi:10.1016/J.AGSY.2016.09.010.
- 26 Karabulut, A. A., E. Crenna, S. Sala, and A. Udias, 2018: A proposal for integration of the ecosystem-
27 water-food-land-energy (EWFLE) nexus concept into life cycle assessment: A synthesis matrix
28 system for food security. *J. Clean. Prod.*, doi:10.1016/j.jclepro.2017.05.092.
- 29 Karar, E., and I. Jacobs-Mata, 2016: Inclusive Governance: The Role of Knowledge in Fulfilling the
30 Obligations of Citizens. *Aquat. Procedia*, **6**, 15–22, doi:10.1016/j.aqpro.2016.06.003.
- 31 Karim, M. R., and A. Thiel, 2017: Role of community based local institution for climate change
32 adaptation in the Teesta riverine area of Bangladesh. *Clim. Risk Manag.*,
33 doi:10.1016/j.crm.2017.06.002.
- 34 Karlberg, L., and Coauthors, 2015: Tackling Complexity: Understanding the Food-Energy-
35 Environment Nexus in Ethiopia’s Lake Tana Sub-basin. *Water Altern.*, **8**. [www.water-](http://www.water-alternatives.org)
36 [alternatives.org](http://www.water-alternatives.org) (Accessed April 13, 2019).
- 37 Karnib, A., 2017: A Quantitative Nexus Approach to Analyse the Interlinkages across the Sustainable
38 Development Goals. *J. Sustain. Dev.*, **10**, 173–180.
- 39 Karpouzoglou, T., A. Dewulf, and J. Clark, 2016: Advancing adaptive governance of social-
40 ecological systems through theoretical multiplicity. *Environ. Sci. Policy*, **57**, 1–9,
41 doi:10.1016/j.envsci.2015.11.011.
- 42 Kaspersen, R. E., 2012: Coping with deep uncertainty: Challenges for environmental assessment and
43 decision-making. *Uncertainty and Risk: Multidisciplinary Perspectives*.
- 44 Kates, R. W., W. R. Travis, and T. J. Wilbanks, 2012: Transformational adaptation when incremental
45 adaptations to climate change are insufficient. *Proc. Natl Acad. Sci. Usa*, **109**, 7156–7161.
- 46 Kaufmann, D., 2009: Governance Matters VIII Aggregate and Individual Governance Indicators.
47 *World Bank Policy Res. Work. Pap. No. 4978*, doi:10.1080/713701075.

- 1 Kaval, P., J. Loomis, and A. Seidl, 2007: Willingness-to-pay for prescribed fire in the Colorado
2 (USA) wildland urban interface. *For. Policy Econ.*, **9**, 928–937.
- 3 Keenan, R. J., 2015: Climate change impacts and adaptation in forest management: a review. *Ann.*
4 *For. Sci.*, doi:10.1007/s13595-014-0446-5.
- 5 Kelkar, N., 2016: Digging Our Rivers' Graves. *Dams,Rivers People Newsl.*, **14**, 1–6.
6 <https://sandrp.in/2016/02/19/digging-our-rivers-graves/>.
- 7 Kelley, C. P., S. Mohtadi, M. A. Cane, R. Seager, and Y. Kushnir, 2015: Climate change in the Fertile
8 Crescent and implications of the recent Syrian drought. *Proc. Natl. Acad. Sci.*, **112**, 3241–3246,
9 doi:10.1073/pnas.1421533112. <http://www.pnas.org/lookup/doi/10.1073/pnas.1421533112>.
- 10 Kemp, R., S. Parto, and R. Gibson, 2005: Governance for sustainable development: moving from
11 theory to practice. *Internaional J. Sustain. Dev.*, doi:10.1504/IJSD.2005.007372.
- 12 Kern, F., and M. Howlett, 2009: Implementing transition management as policy reforms: A case study
13 of the Dutch energy sector. *Policy Sci.*, **42**, 391–408, doi:10.1007/s11077-009-9099-x.
- 14 Kerr, S., and A. Sweet, 2008: Inclusion of agriculture into a domestic emissions trading scheme: New
15 Zealand's experience to date. *Farm Policy J.*, **4**.
- 16 Kesternich, M., C. Reif, and D. Rübhelke, 2017: Recent Trends in Behavioral Environmental
17 Economics. *Environ. Resour. Econ.*, **67**, 403–411, doi:10.1007/s10640-017-0162-3.
- 18 Khan, M. R., and J. T. Roberts, 2013: Adaptation and international climate policy. *Wiley Interdiscip.*
19 *Rev. Clim. Chang.*, **4**, 171–189, doi:10.1002/wcc.212.
- 20 Khandelwal, M., M. E. Hill, P. Greenough, J. Anthony, M. Quill, M. Linderman, and H. S.
21 Udaykumar, 2017: Why Have Improved Cook-Stove Initiatives in India Failed? *World Dev.*, **92**,
22 13–27, doi:10.1016/j.worlddev.2016.11.006. <http://dx.doi.org/10.1016/j.worlddev.2016.11.006>
23 (Accessed April 15, 2019).
- 24 Kibler, K. M., and D. D. Tullios, 2013: Cumulative biophysical impact of small and large hydropower
25 development in Nu River, China. *Water Resour. Res.*, **49**, 3104–3118.
- 26 Kiendrebeogo, Y., K. Assimaidou, and A. Tall, 2017: Social protection for poverty reduction in times
27 of crisis. *J. Policy Model.*, **39**, 1163–1183, doi:10.1016/j.jpolmod.2017.09.003.
- 28 Kim, K., T. Park, S. Bang, and H. Kim, 2017: Real Options-based framework for hydropower plant
29 adaptation to climate change. *J. Manag. Eng.*, **33**.
- 30 Kimmerer, R. W., 2000: Native Knowledge for Native Ecosystems. *J. For.*, **98**, 4–9,
31 doi:10.1093/jof/98.8.4. <https://academic.oup.com/jof/article/98/8/4/4614233> (Accessed April 7,
32 2019).
- 33 Kiruki, H. M., E. H. van der Zanden, Ž. Malek, and P. H. Verburg, 2017a: Land Cover Change and
34 Woodland Degradation in a Charcoal Producing Semi-Arid Area in Kenya. *L. Degrad. Dev.*, **28**,
35 472–481, doi:10.1002/ldr.2545. <http://doi.wiley.com/10.1002/ldr.2545> (Accessed May 23,
36 2018).
- 37 —, —, —, and —, 2017b: Land Cover Change and Woodland Degradation in a Charcoal
38 Producing Semi-Arid Area in Kenya. *L. Degrad. Dev.*, **28**, 472–481, doi:10.1002/ldr.2545.
39 <http://doi.wiley.com/10.1002/ldr.2545> (Accessed May 18, 2018).
- 40 Kissinger, G., M. Herold, and V. De Sy, 2012: Drivers of Deforestation and Forest Degradation: A
41 Synthesis Report for REDD+ Policymakers. 48.
- 42 —, A. Gupta, I. Mulder, and N. Unterstell, 2019: Climate financing needs in the land sector under
43 the Paris Agreement: An assessment of developing country perspectives. *Land use policy*, **83**,
44 256–269, doi:10.1016/j.landusepol.2019.02.007.
45 <https://linkinghub.elsevier.com/retrieve/pii/S0264837717313728>.
- 46 Kivimaa, P., and F. Kern, 2016: Creative destruction or mere niche support? Innovation policy mixes

- 1 for sustainability transitions. *Res. Policy*, **45**, 205–217, doi:10.1016/j.respol.2015.09.008.
- 2 ———, M. Hildén, D. Huitema, A. Jordan, and J. Newig, 2017a: Experiments in climate governance –
3 A systematic review of research on energy and built environment transitions. *J. Clean. Prod.*,
4 doi:10.1016/j.jclepro.2017.01.027.
- 5 ———, H. L. Kangas, and D. Lazarevic, 2017b: Client-oriented evaluation of “creative destruction” in
6 policy mixes: Finnish policies on building energy efficiency transition. *Energy Research and*
7 *Social Science*.
- 8 Klein, R.J.T., G.F. Midgley, B.L. Preston, M. Alam, F.G.H. Berkhout, K. D., and M. . Shaw, 2014:
9 Adaptation opportunities, constraints, and limits. *In: Climate Change 2014: Impacts,*
10 *Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working*
11 *Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change,*
12 *Field, C.B., V.R. Barros, D.J. Dokken, K.J. Mach, M.D. Mastrandrea, T.E. Bilir, M. Chatterjee,*
13 *K.L. Ebi, Y.O. Estrada, R.C. Genova, B. Girma, E.S. Kissel, A.N. Levy, S. MacCracken, P..*
14 *Mastreanda, and L.. White, Eds., Cambridge University Press, 899–943.*
- 15 Klein, R. J. T., and Coauthors, 2015: Adaptation opportunities, constraints, and limits. *Climate*
16 *Change 2014 Impacts, Adaptation and Vulnerability: Part A: Global and Sectoral Aspects*, 899–
17 944.
- 18 Kleindorfer, P. R., H. Kunreuther, and C. Ou-Yang, 2012: Single-year and multi-year insurance
19 policies in a competitive market. *J. Risk Uncertain.*, **45**, 51–78, doi:10.1007/s11166-012-9148-2.
20 <http://link.springer.com/10.1007/s11166-012-9148-2> (Accessed April 3, 2019).
- 21 Kline, K. L., and Coauthors, 2017: Reconciling food security and bioenergy: priorities for action.
22 *GCB Bioenergy*, **9**, 557–576, doi:10.1111/gcbb.12366. <http://doi.wiley.com/10.1111/gcbb.12366>
23 (Accessed May 23, 2018).
- 24 Knoke, T., K. Messerer, and C. Paul, 2017: The Role of Economic Diversification in Forest
25 Ecosystem Management. *Curr. For. Reports*, **3**, 93–106, doi:10.1007/s40725-017-0054-3.
26 <https://doi.org/10.1007/s40725-017-0054-3>.
- 27 Knook, J., V. Eory, M. Brander, and D. Moran, 2018: Evaluation of farmer participatory extension
28 programmes. *J. Agric. Educ. Ext.*, **24**, 309–325, doi:10.1080/1389224X.2018.1466717.
29 <https://doi.org/10.1080/1389224X.2018.1466717>.
- 30 Knorr, W., A. Arneth, and L. Jiang, 2016a: Demographic controls of future global fire risk. *Nat. Clim.*
31 *Chang.*, **6**, 781–785, doi:10.1038/nclimate2999. <http://www.nature.com/articles/nclimate2999>
32 (Accessed April 12, 2019).
- 33 ———, ———, and ———, 2016b: Demographic controls of future global fire risk. *Nat. Clim. Chang.*, **6**,
34 781–785, doi:10.1038/nclimate2999.
- 35 Koechlin, L., J. Quan, and H. Mulukutla, 2016: *Tackling corruption in land governance.*
36 www.landportal.info/partners/legend (Accessed April 13, 2019).
- 37 Koh, I., S. Kim, and D. Lee, 2010: Effects of bibosoop plantation on wind speed, humidity, and
38 evaporation in a traditional agricultural landscape of Korea: Field measurements and modeling.
39 *Agric. Ecosyst. Environ.*, **135**, 294–303, doi:10.1016/J.AGEE.2009.10.008.
40 <https://www.sciencedirect.com/science/article/pii/S0167880909003119> (Accessed April 7,
41 2019).
- 42 Koizumi, T., 2014: Biofuels and food security. *SpringerBriefs in Applied Sciences and Technology*.
- 43 Kolstad, C., and Coauthors, 2014: Social , Economic and Ethical Concepts and Methods. *Climate*
44 *Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth*
45 *Assessment Report of the Intergovernmental Panel on Climate Change*, O. Edenhofer et al.,
46 Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- 47 Kompas, T., V. H. Pham, and T. N. Che, 2018: The Effects of Climate Change on GDP by Country
48 and the Global Economic Gains From Complying With the Paris Climate Accord. *Earth’s*

- 1 *Futur.*, **6**, 1153–1173, doi:10.1029/2018EF000922. <https://doi.org/10.1029/2018EF000922>.
- 2 Koohafkan, P., and M. A. Altieri, 2011: *Globally Important Agricultural Heritage Systems A Legacy*
3 *for the Future GIAHS Globally Important Agricultural Heritage Systems*. Rome,
4 http://www.fao.org/fileadmin/templates/giahs/PDF/GIAHS_Booklet_EN_WEB2011.pdf
5 (Accessed April 15, 2019).
- 6 Koontz, T. M., D. Gupta, P. Mudliar, and P. Ranjan, 2015: Adaptive institutions in social-ecological
7 systems governance: A synthesis framework. *Environ. Sci. Policy*,
8 doi:10.1016/j.envsci.2015.01.003.
- 9 Van Koppen, B., L. Hope, and W. Colenbrander, 2013a: *Gender Aspects of Small-scale Private*
10 *Irrigation in Africa, IWMI Working Paper*. www.iwmi.org (Accessed April 2, 2019).
- 11 —, —, and —, 2013b: *Gender Aspects of Small-scale Private Irrigation in Africa, IWMI*
12 *Working Paper*.
- 13 Kousky, C., and R. Cooke, 2012: Explaining the Failure to Insure Catastrophic Risks. *Geneva Pap.*
14 *Risk Insur. - Issues Pract.*, **37**, 206–227, doi:10.1057/gpp.2012.14.
15 <http://link.springer.com/10.1057/gpp.2012.14> (Accessed April 3, 2019).
- 16 —, E. O. Michel-Kerjan, and P. A. Raschky, 2018a: Does federal disaster assistance crowd out
17 flood insurance? *J. Environ. Econ. Manage.*, **87**, 150–164, doi:10.1016/j.jeem.2017.05.010.
18 <https://linkinghub.elsevier.com/retrieve/pii/S0095069617303479> (Accessed February 26, 2019).
- 19 —, —, —, J. Linnerooth-Bayer, and S. Hochrainer-Stigler, 2018b: Does federal disaster
20 assistance crowd out flood insurance? *J. Environ. Econ. Manage.*, **133**,
21 doi:10.1016/j.jeem.2017.05.010.
- 22 Krätli, S., 2008: Cattle Breeding, Complexity and Mobility in a Structurally Unpredictable
23 Environment: The WoDaaBe Herders of Niger. *Nomad. People.*, **12**, 11–41,
24 doi:10.3167/np.2008.120102. [http://openurl.ingenta.com/content/xref?genre=article&issn=0822-
25 7942&volume=12&issue=1&spage=11](http://openurl.ingenta.com/content/xref?genre=article&issn=0822-7942&volume=12&issue=1&spage=11) (Accessed April 7, 2019).
- 26 —, and N. Schareika, 2010: Living Off Uncertainty: The Intelligent Animal Production of Dryland
27 Pastoralists. *Eur. J. Dev. Res.*, **22**, 605–622, doi:10.1057/ejdr.2010.41.
28 <http://link.springer.com/10.1057/ejdr.2010.41> (Accessed April 7, 2019).
- 29 Krause, A., and Coauthors, 2017: Global consequences of afforestation and bioenergy cultivation on
30 ecosystem service indicators. *Biogeosciences*, **14**, 4829–4850, doi:10.5194/bg-14-4829-2017.
31 <https://www.biogeosciences.net/14/4829/2017/> (Accessed December 8, 2017).
- 32 —, and Coauthors, 2018: Large uncertainty in carbon uptake potential of land-based climate-
33 change mitigation efforts. *Glob. Chang. Biol.*, **24**, 3025–3038, doi:10.1111/gcb.14144.
- 34 Krishnaswamy, J., R. John, and S. Joseph, 2014: Consistent response of vegetation dynamics to recent
35 climate change in tropical mountain regions. *Glob. Chang. Biol.*, **20**, 203–215.
- 36 Kristjanson, P., A. Waters-Bayer, N. Johnson, A. Tipilda, J. Njuki, I. Baltenweck, D. Grace, and S.
37 MacMillan, 2014: Livestock and women’s livelihoods. *Gender in Agriculture: Closing the*
38 *Knowledge Gap*, 209–234.
- 39 Krug, J. H. A., 2018: Accounting of GHG emissions and removals from forest management: a long
40 road from Kyoto to Paris. *Carbon Balance Manag.*, **13**, 1, doi:10.1186/s13021-017-0089-6.
41 <https://cbmjournal.springeropen.com/articles/10.1186/s13021-017-0089-6> (Accessed May 22,
42 2018).
- 43 Kuhfuss, L., R. Préget, S. Thoyer, N. Hanley, P. Le Coent, and M. Désolé, 2016: Nudges, social
44 norms, and permanence in agri-environmental schemes. *Land Econ.*, **92**, 641–655,
45 doi:10.3368/le.92.4.641. [https://www.scopus.com/inward/record.uri?eid=2-s2.0-
46 84995695081&doi=10.3368%2Fle.92.4.641&partnerID=40&md5=7cd4c76f7009a72ccd224443
47 782bb4f6](https://www.scopus.com/inward/record.uri?eid=2-s2.0-84995695081&doi=10.3368%2Fle.92.4.641&partnerID=40&md5=7cd4c76f7009a72ccd224443782bb4f6).

- 1 Kullenberg, C., and D. Kasperowski, 2016: What Is Citizen Science? – A Scientometric Meta-
2 Analysis. *PLoS One*, **11**, e0147152, doi:10.1371/journal.pone.0147152.
3 <http://dx.plos.org/10.1371/journal.pone.0147152> (Accessed December 11, 2017).
- 4 Kumar Nath, T., T. Kumar Baul, M. M. Rahman, M. T. Islam, and M. Harun-Or-Rashid, 2013a:
5 *TRADITIONAL BIOMASS FUEL CONSUMPTION BY RURAL HOUSEHOLDS IN*
6 *DEGRADED SAL (SHOREA ROBUSTA) FOREST AREAS OF BANGLADESH*. 537-544 pp.
7 www.ijetae.com (Accessed April 14, 2019).
- 8 —, —, —, —, and —, 2013b: *TRADITIONAL BIOMASS FUEL CONSUMPTION BY*
9 *RURAL HOUSEHOLDS IN DEGRADED SAL (SHOREA ROBUSTA) FOREST AREAS OF*
10 *BANGLADESH*. 537-544 pp.
- 11 Kumpula, T., A. Pajunen, E. Kaarlejärvi, B. C. Forbes, and F. Stammer, 2011: Land use and land
12 cover change in Arctic Russia: Ecological and social implications of industrial development.
13 *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2010.12.010.
- 14 Kundzewicz, Z. W., 2002: Non-structural Flood Protection and Sustainability. *Water Int.*, **27**, 3–13,
15 doi:10.1080/02508060208686972.
- 16 Kunreuther, H., and R. Lyster, 2016: The role of public and private insurance in reducing losses from
17 extreme weather events and disasters'. *Asia Pacific J. Environ. Law*, **19**, 29–54.
- 18 Kunreuther, H., and Coauthors, 2014: Integrated Risk and Uncertainty Assessment of Climate Change
19 Response Policies. *Climate Change 2014: Mitigation of Climate Change. Contribution of*
20 *Working Group III to the Fifth Assessment Report of the Intergovernmental Panel on Climate*
21 *Change*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge, United Kingdom
22 and New York, NY, USA.
- 23 Kuriakose, A. T., R. Heltberg, W. Wiseman, C. Costella, R. Cipryk, and S. Cornelius, 2012: Climate-
24 Responsive Social Protection Climate - responsive Social Protection. *Strategy*.
- 25 Kurian, M., 2017: The water-energy-food nexus: Trade-offs, thresholds and transdisciplinary
26 approaches to sustainable development. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2016.11.006.
- 27 Kust, G., O. Andreeva, and A. Cowie, 2017: Land Degradation Neutrality: Concept development,
28 practical applications and assessment. *J. Environ. Manage.*, **195**, 16–24,
29 doi:10.1016/j.jenvman.2016.10.043.
- 30 Kuzdas, C., A. Wiek, B. Warner, R. Vignola, and R. Morataya, 2015: Integrated and participatory
31 analysis of water governance regimes: The case of the Costa Rican dry tropics. *World Dev.*, **66**,
32 254–266, doi:10.1016/j.worlddev.2014.08.018.
- 33 Kwakkel, J. H., M. Haasnoot, and W. E. Walker, 2016: Comparing Robust Decision-Making and
34 Dynamic Adaptive Policy Pathways for model-based decision support under deep uncertainty.
35 *Environ. Model. Softw.*, **86**, 168–183, doi:10.1016/j.envsoft.2016.09.017.
- 36 Kweka, E. J., E. E. Kimaro, and S. Munga, 2016: Effect of Deforestation and Land Use Changes on
37 Mosquito Productivity and Development in Western Kenya Highlands: Implication for Malaria
38 Risk. *Front. public Heal.*, **4**, 238, doi:10.3389/fpubh.2016.00238.
39 <http://www.ncbi.nlm.nih.gov/pubmed/27833907> (Accessed December 29, 2017).
- 40 Laakso, S., A. Berg, and M. Annala, 2017: Dynamics of experimental governance: A meta-study of
41 functions and uses of climate governance experiments. *J. Clean. Prod.*,
42 doi:10.1016/j.jclepro.2017.04.140.
- 43 Laczko, F., and E. Piguet, 2014: Regional Perspectives on Migration, the Environment and Climate
44 Change. *People on the Move in an Changing Climate: The Regional Impact of Environmental*
45 *Change on Migration*.
- 46 Lahiff, E., 2015: The great African land grab? Agricultural investments and the global food system. *J.*
47 *Peasant Stud.*, **42**, 239–242, doi:10.1080/03066150.2014.978141.

- 1 Lal, P. N., T. Mitchell, P. Aldunce, H. Auld, R. Mechler, A. Miyan, L. E. Romano, and S. Zakaria,
2 2012: National Systems for Managing the Risks from Climate Extremes and Disasters.
3 *Managing the Risks of Extreme Events and Disasters to Advance Climate Change Adaptation. A*
4 *Special Report of Working Groups I and II of the Intergovernmental Panel on Climate Change*
5 *(IPCC)*, C.B. Field et al., Eds., Cambridge University Press, Cambridge, UK, and New York,
6 NY, USA, 339–392.
- 7 Lal, R., 2013: Food security in a changing climate. *Ecohydrol. Hydrobiol.*,
8 doi:10.1016/j.ecohyd.2013.03.006.
- 9 Lam, P. T. I., and A. O. K. Law, 2016: Crowdfunding for renewable and sustainable energy projects:
10 An exploratory case study approach. *Renew. Sustain. Energy Rev.*, **60**, 11–20,
11 doi:10.1016/j.rser.2016.01.046.
- 12 Lam, Q. D., B. Schmalz, and N. Fohrer, 2011: The impact of agricultural Best Management Practices
13 on water quality in a North German lowland catchment. *Environ. Monit. Assess.*,
14 doi:10.1007/s10661-011-1926-9.
- 15 Lamarque, P., A. Artaux, C. Barnaud, L. Dobremez, B. Nettier, and S. Lavorel, 2013: Taking into
16 account farmers' decision making to map fine-scale land management adaptation to climate and
17 socio-economic scenarios. *Landsc. Urban Plan.*, **119**, 147–157,
18 doi:10.1016/j.landurbplan.2013.07.012. [https://ac-els-cdn-](https://ac-els-cdn-com.ezp.sub.su.se/S0169204613001503/1-s2.0-S0169204613001503-main.pdf?_tid=258dc34c-e703-11e7-8467-00000aab0f01&acdnat=1513938816_91bf3d0d1be78804d5c5b2418e4336c2)
19 [com.ezp.sub.su.se/S0169204613001503/1-s2.0-S0169204613001503-main.pdf?_tid=258dc34c-](https://ac-els-cdn-com.ezp.sub.su.se/S0169204613001503/1-s2.0-S0169204613001503-main.pdf?_tid=258dc34c-e703-11e7-8467-00000aab0f01&acdnat=1513938816_91bf3d0d1be78804d5c5b2418e4336c2)
20 [e703-11e7-8467-00000aab0f01&acdnat=1513938816_91bf3d0d1be78804d5c5b2418e4336c2](https://ac-els-cdn-com.ezp.sub.su.se/S0169204613001503/1-s2.0-S0169204613001503-main.pdf?_tid=258dc34c-e703-11e7-8467-00000aab0f01&acdnat=1513938816_91bf3d0d1be78804d5c5b2418e4336c2)
21 (Accessed December 22, 2017).
- 22 Lambin, E. F., and Coauthors, 2001: The causes of land-use and land-cover change: moving beyond
23 the myths. *Glob. Environ. Chang.*, **11**, 261–269, doi:10.1016/S0959-3780(01)00007-3.
24 <http://linkinghub.elsevier.com/retrieve/pii/S0959378001000073> (Accessed May 22, 2018).
- 25 ———, and Coauthors, 2014: Effectiveness and synergies of policy instruments for land use governance
26 in tropical regions. *Glob. Environ. Chang.*, **28**, 129–140,
27 doi:10.1016/J.GLOENVCHA.2014.06.007. [https://www.sciencedirect-](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0959378014001125)
28 [com.ezp.sub.su.se/science/article/pii/S0959378014001125](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0959378014001125) (Accessed May 23, 2018).
- 29 Lane, C., and R. Moorehead, 1995: New directions in rangeland and resource tenure and policy.
30 *Living with Uncertainty: New directions in pastoral Development in Africa.*, I. Scoones, Ed.,
31 Practical Action Publishing, Rugby, Warwickshire, United Kingdom, 116–133
32 <https://www.developmentbookshelf.com/doi/10.3362/9781780445335.007> (Accessed November
33 3, 2018).
- 34 Lane, C. R., 1998: *Custodians of the commons: pastoral land tenure in East and West Africa.*
35 Earthscan, 238 pp.
36 [https://books.google.co.uk/books?hl=en&lr=&id=LB9pAwAAQBAJ&oi=fnd&pg=PP1&dq=lan](https://books.google.co.uk/books?hl=en&lr=&id=LB9pAwAAQBAJ&oi=fnd&pg=PP1&dq=lane+custodians+of+the+commons&ots=XCsnJKXH_c&sig=KRkhlsb0ix38asdEnSYQq_wyqz4#v=onepage&q=lane+custodians+of+the+commons&f=false)
37 [e+custodians+of+the+commons&ots=XCsnJKXH_c&sig=KRkhlsb0ix38asdEnSYQq_wyqz4#v](https://books.google.co.uk/books?hl=en&lr=&id=LB9pAwAAQBAJ&oi=fnd&pg=PP1&dq=lane+custodians+of+the+commons&ots=XCsnJKXH_c&sig=KRkhlsb0ix38asdEnSYQq_wyqz4#v=onepage&q=lane+custodians+of+the+commons&f=false)
38 [=onepage&q=lane custodians of the commons&f=false](https://books.google.co.uk/books?hl=en&lr=&id=LB9pAwAAQBAJ&oi=fnd&pg=PP1&dq=lane+custodians+of+the+commons&ots=XCsnJKXH_c&sig=KRkhlsb0ix38asdEnSYQq_wyqz4#v=onepage&q=lane+custodians+of+the+commons&f=false) (Accessed May 22, 2018).
- 39 Lane, R., and R. McNaught, 2009: Building gendered approaches to adaptation in the Pacific. *Gend.*
40 *Dev.*, **17**, 67–80, doi:10.1080/13552070802696920.
41 <https://www.tandfonline.com/doi/full/10.1080/13552070802696920> (Accessed April 8, 2019).
- 42 Lange, E., and S. Hehl-Lange, 2011: Citizen participation in the conservation and use of rural
43 landscapes in Britain: The Alport Valley case study. *Landsc. Ecol. Eng.*, **7**, 223–230,
44 doi:10.1007/s11355-010-0115-2.
- 45 Langholtz, M., and Coauthors, 2014: Climate risk management for the U.S. cellulosic biofuels supply
46 chain. *Clim. Risk Manag.*, doi:10.1016/j.crm.2014.05.001.
- 47 Larcom, S., and T. van Gevelt, 2017: Regulating the water-energy-food nexus: Interdependencies,
48 transaction costs and procedural justice. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2017.03.003.
- 49 Larkin, A., J. Kuriakose, M. Sharmina, and K. Anderson, 2018: What if negative emission

- 1 technologies fail at scale? Implications of the Paris Agreement for big emitting nations. *Clim.*
2 *Policy*, **18**, 690–714, doi:10.1080/14693062.2017.1346498.
3 <https://www.tandfonline.com/doi/full/10.1080/14693062.2017.1346498> (Accessed March 30,
4 2019).
- 5 Larsen, P. H., S. Goldsmith, O. Smith, M. L. Wilson, K. Strzepek, P. Chinowsky, and B. Saylor,
6 2008: Estimating future costs for Alaska public infrastructure at risk from climate change. *Glob.*
7 *Environ. Chang.*, **18**, 442–457, doi:10.1016/j.gloenvcha.2008.03.005.
- 8 Larson, A., and J. Pulhin, 2012: Enhancing forest tenure reforms through more responsive regulations.
9 *Conserv. Soc.*, **10**, 103, doi:10.4103/0972-4923.97482.
- 10 Lashley, J. G., and K. Warner, 2015: Evidence of demand for microinsurance for coping and
11 adaptation to weather extremes in the Caribbean. *Clim. Change*, doi:10.1007/s10584-013-0922-
12 1.
- 13 Laube, W., B. Schraven, and M. Awo, 2012: Smallholder adaptation to climate change: Dynamics and
14 limits in Northern Ghana. *Clim. Change*, doi:10.1007/s10584-011-0199-1.
- 15 Lavell, A., and Coauthors, 2012: Climate change: New dimensions in disaster risk, exposure,
16 vulnerability, and resilience. *Managing the Risks of Extreme Events and Disasters to Advance*
17 *Climate Change Adaptation: Special Report of the Intergovernmental Panel on Climate Change*,
18 Vol. 9781107025 of, 25–64.
- 19 Lavers, T., 2012: ‘Land grab’ as development strategy? The political economy of agricultural
20 investment in Ethiopia. *J. Peasant Stud.*, **39**, 105–132, doi:10.1080/03066150.2011.652091.
- 21 Lawrence, J., R. Bell, P. Blackett, S. Stephens, and S. Allan, 2018: National guidance for adapting to
22 coastal hazards and sea-level rise: Anticipating change, when and how to change pathway.
23 *Environ. Sci. Policy*, doi:10.1016/j.envsci.2018.01.012.
- 24 ———, R. Bell, and A. Stroombergen, 2019: A hybrid process to address uncertainty and changing
25 climate risk in coastal areas using Dynamic adaptive pathways planning, multi-criteria decision
26 analysis & Real options analysis: A New Zealand application. *Sustain.*, **11**, 1–18,
27 doi:10.3390/su11020406.
- 28 Layke, C., 2009: Measuring nature’s benefits: a preliminary roadmap for improving ecosystem
29 service indicators. *World Resour. Inst. Washingt.*,.
- 30 Lazo, J. K., A. Bostrom, R. Morss, J. Demuth, and H. Lazrus, *Communicating Hurricane Warnings:*
31 *Factors Affecting Protective Behavior.* [https://cdn1.sph.harvard.edu/wp-](https://cdn1.sph.harvard.edu/wp-content/uploads/sites/1273/2014/02/Risk-Perception-Lazo-et-al.pdf)
32 [content/uploads/sites/1273/2014/02/Risk-Perception-Lazo-et-al.pdf](https://cdn1.sph.harvard.edu/wp-content/uploads/sites/1273/2014/02/Risk-Perception-Lazo-et-al.pdf) (Accessed April 3, 2019).
- 33 Leach, G., 1992: The energy transition. *Energy Policy*, **20**, 116–123, doi:10.1016/0301-
34 4215(92)90105-B. [https://www.sciencedirect-](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/030142159290105B)
35 [com.ezp.sub.su.se/science/article/pii/030142159290105B](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/030142159290105B) (Accessed April 14, 2019).
- 36 Ledec, George C.; Rapp, Kennan W.; Aiello, R. G., 2011: *Greening the wind: environmental and*
37 *social considerations for wind power development.* World Bank., Washington, DC,
38 [http://documents.worldbank.org/curated/en/239851468089382658/Greening-the-wind-](http://documents.worldbank.org/curated/en/239851468089382658/Greening-the-wind-environmental-and-social-considerations-for-wind-power-development)
39 [environmental-and-social-considerations-for-wind-power-development](http://documents.worldbank.org/curated/en/239851468089382658/Greening-the-wind-environmental-and-social-considerations-for-wind-power-development).
- 40 Lee, C. M., and M. Lazarus, 2013: Bioenergy projects and sustainable development: which project
41 types offer the greatest benefits? *Clim. Dev.*, **5**, 305–317, doi:10.1080/17565529.2013.812951.
42 <http://www.tandfonline.com/doi/abs/10.1080/17565529.2013.812951> (Accessed May 22, 2018).
- 43 Lee, M., C. Armeni, J. de Cendra, S. Chaytor, S. Lock, M. Maslin, C. Redgwell, and Y. Rydin, 2013:
44 Public participation and climate change infrastructure. *J. Environ. Law*, **25**, 33–62,
45 doi:10.1093/jel/eqs027.
- 46 Lele, U., M. Klousia-Marquis, and S. Goswami, 2013: Good Governance for Food, Water and Energy
47 Security. *Aquat. Procedia*, doi:10.1016/j.aqpro.2013.07.005.

- 1 Lelieveld, J., C. Barlas, D. Giannadaki, and A. Pozzer, 2013: Model calculated global, regional and
2 megacity premature mortality due to air pollution. *Atmos. Chem. Phys.*, **13**, 7023–7037,
3 doi:10.5194/acp-13-7023-2013.
- 4 Lemann, A. B., 2018: Stronger Than the Storm: Disaster Law in a Defiant Age. *78 La. L. Rev.*,
5 <http://scholarship.law.marquette.edu/facpubhttp://scholarship.law.marquette.edu/facpub/687>
6 (Accessed October 3, 2018).
- 7 Lempert, R., 2013: Scenarios that illuminate vulnerabilities and robust responses. *Clim. Change*, **117**,
8 627–646, doi:10.1007/s10584-012-0574-6.
- 9 Lempert, R. J., and M. E. Schlesinger, 2000: Robust strategies for abating climate change. *Clim.*
10 *Change*, **45**, 387–401.
- 11 Lenderink, G., and E. van Meijgaard, 2008: Increase in hourly precipitation extremes beyond
12 expectations from temperature changes. *Nat. Geosci.*, **1**, 511–514, doi:10.1038/ngeo262.
13 <http://www.nature.com/articles/ngeo262> (Accessed April 14, 2019).
- 14 Leonard, S., M. Parsons, K. Olawsky, and F. Kofod, 2013: The role of culture and traditional
15 knowledge in climate change adaptation: Insights from East Kimberley, Australia. *Glob.*
16 *Environ. Chang.*, **23**, 623–632, doi:10.1016/J.GLOENVCHA.2013.02.012.
17 <https://www.sciencedirect.com/science/article/abs/pii/S0959378013000423> (Accessed April 8,
18 2019).
- 19 Leys, A. J., and J. K. Vanclay, 2011: Social learning: A knowledge and capacity building approach
20 for adaptive co-management of contested landscapes. *Land use policy*, **28**, 574–584,
21 doi:10.1016/j.landusepol.2010.11.006.
- 22 Li, Z., and H. Fang, 2016a: Impacts of climate change on water erosion: A review. *Earth-Science*
23 *Rev.*, **163**, 94–117, doi:10.1016/J.EARSCIREV.2016.10.004.
24 <https://www.sciencedirect.com/science/article/pii/S0012825216303555> (Accessed April 14,
25 2019).
- 26 ———, and ———, 2016b: Impacts of climate change on water erosion: A review. *Earth-Science Rev.*,
27 **163**, 94–117, doi:10.1016/J.EARSCIREV.2016.10.004.
- 28 Lilleør, H. B., and K. Van den Broeck, 2011: Economic drivers of migration and climate change in
29 LDCs. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2011.09.002.
- 30 Lin, B., and X. Li, 2011: The effect of carbon tax on per capita CO₂ emissions. *Energy Policy*, **39**,
31 5137–5146, doi:10.1016/j.enpol.2011.05.050.
- 32 Lin, Y.-P., D. Deng, W.-C. Lin, R. Lemmens, N. D. Crossman, K. Henle, and D. S. Schmeller, 2015:
33 Uncertainty analysis of crowd-sourced and professionally collected field data used in species
34 distribution models of Taiwanese moths. *Biol. Conserv.*, **181**, 102–110,
35 doi:10.1016/J.BIOCON.2014.11.012.
36 <https://www.sciencedirect.com/science/article/pii/S0006320714004376> (Accessed October 2,
37 2018).
- 38 Lindner, M., and Coauthors, 2010: Climate change impacts, adaptive capacity, and vulnerability of
39 European forest ecosystems. *For. Ecol. Manage.*, **259**, 698–709,
40 doi:10.1016/J.FORECO.2009.09.023.
41 <https://www.sciencedirect.com/science/article/pii/S0378112709006604> (Accessed April 14,
42 2019).
- 43 Liniger, H., N. Harari, G. van Lynden, R. Fleiner, J. de Leeuw, Z. Bai, and W. Critchley, 2019:
44 Achieving land degradation neutrality: The role of SLM knowledge in evidence-based decision-
45 making. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2019.01.001.
- 46 Linnerooth-bayer, J., S. Surminski, L. M. Bouwer, I. Noy, and R. Mechler, 2018: Insurance as a
47 Response to Loss and Damage? *Loss and Damage from Climate Change: Concepts, Methods*
48 *and Policy Options*, R. Mechler, L.M. Bouwer, T. Schinko, S. Surminski, and J. Linnerooth-

- 1 bayer, Eds., Springer.
- 2 —, —, —, —, and —, 2019: Insurance as a Response to Loss and Damage? *Loss and*
3 *Damage from Climate Change: Concepts, Methods and Policy Options*, Springer International
4 Publishing, Berlin.
- 5 Linnerooth-Bayer, J., and S. Hochrainer-Stigler, 2015: Financial instruments for disaster risk
6 management and climate change adaptation. *Clim. Change*, **133**, 85–100, doi:10.1007/s10584-
7 013-1035-6.
- 8 Lipper, L., and Coauthors, 2014a: Climate-smart agriculture for food security. *Nat. Clim. Chang.*, **4**,
9 1068–1072, doi:10.1038/nclimate2437.
- 10 Lipper, L., and Coauthors, 2014b: Climate-smart agriculture for food security. *Nat. Clim. Chang.*,
11 doi:10.1038/nclimate2437.
- 12 Lipper, L., and Coauthors, 2014c: Climate-smart agriculture for food security. *Nat. Clim. Chang.*, **4**,
13 1068–1072, doi:10.1038/nclimate2437. <http://www.nature.com/articles/nclimate2437> (Accessed
14 May 18, 2018).
- 15 Little, P. D., H. Mahmoud, and D. L. Coppock, 2001: When deserts flood: Risk management and
16 climatic processes among East African pastoralists. *Clim. Res.*, doi:10.3354/cr019149.
- 17 Liu, J., and Coauthors, 2017: Challenges in operationalizing the water–energy–food nexus. *Hydrol.*
18 *Sci. J.*, **62**, 1714–1720, doi:10.1080/02626667.2017.1353695.
- 19 Liu, P., and N. Ravenscroft, 2017: Collective action in implementing top-down land policy: The case
20 of Chengdu, China. *Land use policy*, **65**, 45–52, doi:10.1016/J.LANDUSEPOL.2017.03.031.
21 <https://www.sciencedirect.com/science/article/pii/S0264837716309140> (Accessed May 23,
22 2018).
- 23 Loaiza, T., U. Nehren, and G. Gerold, 2015: REDD+ and incentives: An analysis of income
24 generation in forest-dependent communities of the Yasuní Biosphere Reserve, Ecuador. *Appl.*
25 *Geogr.*, **62**, 225–236, doi:10.1016/J.APGEOG.2015.04.020.
26 <https://www.sciencedirect.com/science/article/abs/pii/S0143622815001034> (Accessed October
27 15, 2018).
- 28 —, M. O. Borja, U. Nehren, and G. Gerold, 2017: Analysis of land management and legal
29 arrangements in the Ecuadorian Northeastern Amazon as preconditions for REDD+
30 implementation. *For. Policy Econ.*, **83**, 19–28, doi:10.1016/J.FORPOL.2017.05.005.
31 <https://www.sciencedirect.com/science/article/pii/S1389934116303938> (Accessed October 15,
32 2018).
- 33 Loarie, S. R., P. B. Duffy, H. Hamilton, G. P. Asner, C. B. Field, and D. D. Ackerly, 2009: The
34 velocity of climate change. *Nature*, doi:10.1038/nature08649.
- 35 Lobell, D. B., U. L. C. Baldos, and T. W. Hertel, 2013: Climate adaptation as mitigation: The case of
36 agricultural investments. *Environ. Res. Lett.*, **8**, doi:10.1088/1748-9326/8/1/015012.
- 37 Locatelli, B., P. Imbach, and S. Wunder, 2014: Synergies and trade-offs between ecosystem services
38 in Costa Rica. *Environ. Conserv.*, **41**, 27–36.
- 39 —, G. Fedele, V. Fayolle, and A. Baglee, 2016: Synergies between adaptation and mitigation in
40 climate change finance. *Int. J. Clim. Chang. Strateg. Manag.*, **8**, 112–128, doi:10.1108/IJCCSM-
41 07-2014-0088.
- 42 Lontzek, T. S., Y. Cai, K. L. Judd, and T. M. Lenton, 2015: Stochastic integrated assessment of
43 climate tipping points indicates the need for strict climate policy. *Nat. Clim. Chang.*,
44 doi:10.1038/nclimate2570.
- 45 Loos, J., A. I. Horcea-Milcu, P. Kirkland, T. Hartel, M. Osváth-Ferencz, and J. Fischer, 2015:
46 Challenges for biodiversity monitoring using citizen science in transitioning social–ecological
47 systems. *J. Nat. Conserv.*, **26**, 45–48, doi:10.1016/j.jnc.2015.05.001.

- 1 López-i-Gelats, F., E. D. G. Fraser, J. F. Morton, and M. G. Rivera-Ferre, 2016: What drives the
2 vulnerability of pastoralists to global environmental change? A qualitative meta-analysis. *Glob.*
3 *Environ. Chang.*, **39**, 258–274, doi:10.1016/J.GLOENVCHA.2016.05.011.
4 <https://www.sciencedirect.com/science/article/pii/S095937801630070X> (Accessed May 22,
5 2018).
- 6 Lotze-Campen, H., and A. Popp, 2012: Agricultural Adaptation Options: Production Technology,
7 Insurance, Trade. *Climate Change, Justice and Sustainability*, Springer Netherlands, Dordrecht,
8 171–178 http://www.springerlink.com/index/10.1007/978-94-007-4540-7_16 (Accessed May
9 16, 2018).
- 10 Louhaichi, M., and A. Tastad, 2010: The Syrian Steppe: Past Trends, Current Status, and Future
11 Priorities. *Rangelands*, **32**, 2–7, doi:10.2307/40588043. <https://www.jstor.org/stable/40588043>
12 (Accessed April 8, 2019).
- 13 Lubis, R. F., R. Delinom, S. Martosuparno, and H. Bakti, 2018: Water-Food Nexus in Citarum
14 Watershed, Indonesia. *IOP Conference Series: Earth and Environmental Science*.
- 15 Luderer, G., R. C. Pietzcker, C. Bertram, E. Kriegler, M. Meinshausen, and O. Edenhofer, 2013:
16 Economic mitigation challenges: how further delay closes the door for achieving climate
17 targets. doi:10.1088/1748-9326/8/3/034033.
- 18 Lundmark, T., and Coauthors, 2014: Potential Roles of Swedish Forestry in the Context of Climate
19 Change Mitigation. *Forests*, **5**, 557–578, doi:10.3390/f5040557. [http://www.mdpi.com/1999-
20 4907/5/4/557](http://www.mdpi.com/1999-4907/5/4/557) (Accessed April 5, 2019).
- 21 Lunt, T., A. W. Jones, W. S. Mulhern, D. P. M. Lezaks, and M. M. Jahn, 2016: Vulnerabilities to
22 agricultural production shocks: An extreme, plausible scenario for assessment of risk for the
23 insurance sector. *Clim. Risk Manag.*, **13**, 1–9, doi:10.1016/j.crm.2016.05.001.
- 24 Lybbert, T. J., and D. A. Sumner, 2012: Agricultural technologies for climate change in developing
25 countries: Policy options for innovation and technology diffusion. *Food Policy*, **37**, 114–123,
26 doi:10.1016/j.foodpol.2011.11.001. <http://www.aatf-africa.org/userfiles/WEMA-brief.pdf>
27 (Accessed October 15, 2018).
- 28 Lynch, A. J., and Coauthors, 2016: The social, economic, and environmental importance of inland
29 fish and fisheries. *Environ. Rev.*, **24**, 115–121.
- 30 MacGregor, S., 2010: ‘Gender and climate change’: from impacts to discourses. *J. Indian Ocean Reg.*,
31 **6**, 223–238, doi:10.1080/19480881.2010.536669.
32 <http://www.tandfonline.com/doi/abs/10.1080/19480881.2010.536669> (Accessed April 2, 2019).
- 33 Macintosh, A. K., 2012: LULUCF in the post-2012 regime: fixing the problems of the past? *Clim.*
34 *Policy*, **12**, 341–355, doi:10.1080/14693062.2011.605711.
35 <http://www.tandfonline.com/doi/abs/10.1080/14693062.2011.605711> (Accessed May 22, 2018).
- 36 Magnan, A., 2014: Avoiding maladaptation to climate change: towards guiding principles.
37 *S.A.P.I.E.N.S.*.
- 38 Magnan, A. K., and Coauthors, 2016: Addressing the risk of maladaptation to climate change. *Wiley*
39 *Interdiscip. Rev. Clim. Chang.*, **7**, 646–665, doi:10.1002/wcc.409.
- 40 Mahul, O., and F. Ghesquiere, 2010: Financial protection of the state against natural disasters: a
41 primer. doi:10.1596/1813-9450-5429.
- 42 Maikhuri, R. K., R. L. Senwal, K. S. Rao, and K. G. Saxena, 1997: Rehabilitation of degraded
43 community lands for sustainable development in Himalaya: A case study in Garhwal Himalaya,
44 India. *Int. J. Sustain. Dev. World Ecol.*, **4**, 192–203, doi:10.1080/13504509709469954.
45 <https://www.tandfonline.com/doi/full/10.1080/13504509709469954> (Accessed April 7, 2019).
- 46 Majumder, M., 2015: *Impact of Urbanization on Water Shortage in Face of Climatic Aberrations*.
47 Springer Singapore, Singapore, <http://link.springer.com/10.1007/978-981-4560-73-3>.

- 1 Makkonen, M., S. Huttunen, E. Primmer, A. Repo, and M. Hildén, 2015: Policy coherence in climate
2 change mitigation: An ecosystem service approach to forests as carbon sinks and bioenergy
3 sources. *For. Policy Econ.*, **50**, 153–162, doi:10.1016/j.forpol.2014.09.003.
4 <http://dx.doi.org/10.1016/j.forpol.2014.09.003> (Accessed April 12, 2019).
- 5 Maldonado, J. K., C. Shearer, R. Bronen, K. Peterson, and H. Lazrus, 2014: The impact of climate
6 change on tribal communities in the US: Displacement, relocation, and human rights. *Climate*
7 *Change and Indigenous Peoples in the United States: Impacts, Experiences and Actions*.
- 8 Mallampalli, V. R., and Coauthors, 2016: Methods for translating narrative scenarios into quantitative
9 assessments of land use change. *Environ. Model. Softw.*, **82**, 7–20,
10 doi:10.1016/J.ENVSOFT.2016.04.011.
11 <http://www.sciencedirect.com/science/article/pii/S1364815216301037> (Accessed December 12,
12 2017).
- 13 Malogdos, F. K., and E. Yujuico, 2015a: Reconciling formal and informal decision-making on
14 ecotourist infrastructure in Sagada, Philippines. *J. Sustain. Tour.*,
15 doi:10.1080/09669582.2015.1049608.
- 16 ———, and ———, 2015b: Reconciling formal and informal decision-making on ecotourist infrastructure
17 in Sagada, Philippines. *J. Sustain. Tour.*, **23**, 1482–1503, doi:10.1080/09669582.2015.1049608.
18 <http://www.tandfonline.com/doi/full/10.1080/09669582.2015.1049608> (Accessed December 22,
19 2017).
- 20 von Maltitz, G. P., G. Henley, M. Ogg, P. C. Samboko, A. Gasparatos, M. Read, F. Engelbrecht, and
21 A. Ahmed, 2018: Institutional arrangements of outgrower sugarcane production in Southern
22 Africa. *Dev. South. Afr.*, 1–23, doi:10.1080/0376835X.2018.1527215.
23 <https://www.tandfonline.com/doi/full/10.1080/0376835X.2018.1527215> (Accessed April 13,
24 2019).
- 25 Maltsoglou, I., and Coauthors, 2014: Combining bioenergy and food security: An approach and rapid
26 appraisal to guide bioenergy policy formulation. *Biomass and Bioenergy*,
27 doi:10.1016/j.biombioe.2015.02.007.
- 28 Manning, P., G. Taylor, and M. E. Hanley, 2015: Bioenergy, Food Production and Biodiversity - An
29 Unlikely Alliance? *GCB Bioenergy*, **7**, 570–576, doi:10.1111/gcbb.12173.
30 <http://doi.wiley.com/10.1111/gcbb.12173> (Accessed May 24, 2018).
- 31 Mantel, S. K., D. A. Hughes, and N. W. J. Muller, 2010: Ecological impacts of small dams on South
32 African rivers Part 1: drivers of change-water quantity and quality. *Water Sa*, **36**, 351–360.
- 33 Mapfumo, P., F. Mtambanengwe, and R. Chikowo, 2016: Building on indigenous knowledge to
34 strengthen the capacity of smallholder farming communities to adapt to climate change and
35 variability in southern Africa. *Clim. Dev.*, **8**, 72–82, doi:10.1080/17565529.2014.998604.
36 <http://www.tandfonline.com/doi/full/10.1080/17565529.2014.998604> (Accessed April 23,
37 2019).
- 38 ———, and Coauthors, 2017: Pathways to transformational change in the face of climate impacts: an
39 analytical framework. *Clim. Dev.*, **9**, doi:10.1080/17565529.2015.1040365.
- 40 Maraseni, T. N., and T. Cadman, 2015: A comparative analysis of global stakeholders' perceptions of
41 the governance quality of the clean development mechanism (CDM) and reducing emissions
42 from deforestation and forest degradation (REDD+). *Int. J. Environ. Stud.*, **72**, 288–304,
43 doi:10.1080/00207233.2014.993569.
44 <http://www.tandfonline.com/doi/full/10.1080/00207233.2014.993569> (Accessed May 22, 2018).
- 45 Marcacci, S., 2018: India Coal Power Is About To Crash: 65% Of Existing Coal Costs More Than
46 New Wind And Solar. *FORBES ENERGY INNOVATION*
47 [https://www.forbes.com/sites/energyinnovation/2018/01/30/india-coal-power-is-about-to-crash-](https://www.forbes.com/sites/energyinnovation/2018/01/30/india-coal-power-is-about-to-crash-65-of-existing-coal-costs-more-than-new-wind-and-solar/#5c53d9344c0f)
48 [65-of-existing-coal-costs-more-than-new-wind-and-solar/#5c53d9344c0f](https://www.forbes.com/sites/energyinnovation/2018/01/30/india-coal-power-is-about-to-crash-65-of-existing-coal-costs-more-than-new-wind-and-solar/#5c53d9344c0f).
- 49 Marchand, P., and Coauthors, 2016: Reserves and trade jointly determine exposure to food supply

- 1 shocks. *Environ. Res. Lett.*, doi:10.1088/1748-9326/11/9/095009.
- 2 Marjanac, S., L. Patton, and J. Thornton, 2017: Acts of god, human influence and litigation. *Nat.*
3 *Geosci.*, **10**, 616–619, doi:10.1038/ngeo3019.
- 4 Markusson, N., D. McLaren, and D. Tyfield, 2018a: Towards a cultural political economy of
5 mitigation deterrence by negative emissions technologies (NETs). *Glob. Sustain.*, **1**, e10,
6 doi:10.1017/sus.2018.10.
7 https://www.cambridge.org/core/product/identifier/S2059479818000108/type/journal_article
8 (Accessed March 30, 2019).
- 9 —, —, and —, 2018b: Towards a cultural political economy of mitigation deterrence by
10 negative emissions technologies (NETs). *Glob. Sustain.*, **1**, e10, doi:10.1017/sus.2018.10.
- 11 Marshall, N. ., S. . Park, W. N. Adger, K. Brown, and S. . Howden, 2012: Transformational capacity
12 and the influence of place and identity. *Environ. Res. Lett.*, **7**.
- 13 Martin, D. R., and K. L. Pope, 2011: Luring anglers to enhance fisheries. *J. Environ. Manage.*, **92**,
14 1409–1413, doi:10.1016/j.jenvman.2010.10.002.
- 15 St. Martin, K., 2009: Toward a Cartography of the Commons: Constituting the Political and
16 Economic Possibilities of Place. *Prof. Geogr.*, **61**, 493–507, doi:10.1080/00330120903143482.
17 <http://www.tandfonline.com/doi/abs/10.1080/00330120903143482> (Accessed April 9, 2019).
- 18 Martin Persson, U., 2015: The Impact of Biofuel Demand on Agricultural Commodity Prices: A
19 Systematic Review. *Advances in Bioenergy: The Sustainability Challenge*.
- 20 Martinez, G., E. Williams, and S. Yu, 2015: The Economics of Health Damage and Adaptation to
21 Climate Change in Europe: A Review of the Conventional and Grey Literature. *Climate*, **3**, 522–
22 541, doi:10.3390/cli3030522. <http://www.mdpi.com/2225-1154/3/3/522/>.
- 23 Martins, M., P. M. Costa, J. Raimundo, C. Vale, A. M. Ferreira, and M. H. Costa, 2012: Impact of
24 remobilized contaminants in *Mytilus edulis* during dredging operations in a harbour area:
25 bioaccumulation and biomarker responses. *Ecotoxicol. Environ. Saf.*, **85**, 96–103.
- 26 Maserà, O. R., R. Bailis, R. Drigo, A. Ghilardi, and I. Ruiz-Mercado, 2015: Environmental Burden of
27 Traditional Bioenergy Use. *Annu. Rev. Environ. Resour.*, **40**, 121–150, doi:10.1146/annurev-
28 environ-102014-021318. [http://www.annualreviews.org/doi/10.1146/annurev-environ-102014-
29 021318](http://www.annualreviews.org/doi/10.1146/annurev-environ-102014-021318) (Accessed November 16, 2018).
- 30 Mathy, S., and O. Blanchard, 2016: Proposal for a poverty-adaptation-mitigation window within the
31 Green Climate Fund. *Clim. Policy*, **16**, 752–767, doi:10.1080/14693062.2015.1050348.
- 32 Matthies, B. D., T. Kalliokoski, T. Ekholm, H. F. Hoen, and L. T. Valsta, 2015: Risk, reward, and
33 payments for ecosystem services: A portfolio approach to ecosystem services and forestland
34 investment. *Ecosyst. Serv.*, doi:10.1016/j.ecoser.2015.08.006.
- 35 Maxwell, D., and K. Wiebe, 1999: Land tenure and food security: Exploring dynamic linkages. *Dev.*
36 *Change*, **30**, 825–849, doi:10.1111/1467-7660.00139.
- 37 Maynard, T., 2015: *Food system shock: The insurance impacts of acute disruption to global food*
38 *supply*.
- 39 Mayor, B., E. López-Gunn, F. I. Villarroja, and E. Montero, 2015: Application of a water–energy–
40 food nexus framework for the Duero river basin in Spain. *Water Int.*,
41 doi:10.1080/02508060.2015.1071512.
- 42 Maystadt, J. F., and O. Ecker, 2014: Extreme weather and civil war: Does drought fuel conflict in
43 Somalia through livestock price shocks? *Am. J. Agric. Econ.*, doi:10.1093/ajae/aau010.
- 44 McCall, M. K., and C. E. Dunn, 2012: Geo-information tools for participatory spatial planning:
45 Fulfilling the criteria for ‘good’ governance? *Geoforum*, **43**, 81–94,
46 doi:10.1016/j.geoforum.2011.07.007.
47 <http://linkinghub.elsevier.com/retrieve/pii/S0016718511001369> (Accessed November 3, 2018).

- 1 McCarter, J., M. C. Gavin, S. Baereleo, and M. Love, 2014: The challenges of maintaining indigenous
2 ecological knowledge. *Ecol. Soc.*, **19**, art39, doi:10.5751/ES-06741-190339.
3 <http://www.ecologyandsociety.org/vol19/iss3/art39/> (Accessed April 15, 2019).
- 4 McClean, C., R. Whiteley, and N. M. Hayes, 2010: *ISO 31000 — The New , Streamlined Risk*
5 *Management Standard*. 1-4 pp.
- 6 McClelland, S. C., C. Arndt, D. R. Gordon, and G. Thoma, 2018: Type and number of environmental
7 impact categories used in livestock life cycle assessment: A systematic review. *Livest. Sci.*, **209**,
8 39–45, doi:<https://doi.org/10.1016/j.livsci.2018.01.008>.
- 9 McCollum, D. L., and Coauthors, 2018: Connecting the sustainable development goals by their
10 energy inter-linkages. *Environ. Res. Lett.*, **13**, 033006, doi:10.1088/1748-9326/aaafe3.
11 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=3/a=033006?key=crossref.403628ea12e248e0377412ee8730f0cc)
12 [9326/13/i=3/a=033006?key=crossref.403628ea12e248e0377412ee8730f0cc](http://stacks.iop.org/1748-9326/13/i=3/a=033006?key=crossref.403628ea12e248e0377412ee8730f0cc) (Accessed
13 November 16, 2018).
- 14 McDermott, C. L., L. C. Irland, and P. Pacheco, 2015: Forest certification and legality initiatives in the
15 Brazilian Amazon: Lessons for effective and equitable forest governance. *For. Policy Econ.*, **50**,
16 134–142, doi:10.1016/j.forpol.2014.05.011. <http://dx.doi.org/10.1016/j.forpol.2014.05.011>
17 (Accessed April 8, 2019).
- 18 McDowell, N. G., and C. D. Allen, 2015a: Darcy’s law predicts widespread forest mortality under
19 climate warming. *Nat. Clim. Chang.*, **5**, 669–672, doi:10.1038/nclimate2641.
20 <http://www.nature.com/doi/10.1038/nclimate2641> (Accessed November 30, 2017).
- 21 —, and —, 2015b: Darcy’s law predicts widespread forest mortality under climate warming. *Nat.*
22 *Clim. Chang.*, **5**, 669–672, doi:10.1038/nclimate2641.
- 23 McGrath, D. G., F. de Castro, C. Fudemma, B. D. de Amaral, and J. Calabria, 1993: Fisheries and the
24 evolution of resource management on the lower Amazon floodplain. *Hum. Ecol.*, **21**, 167–195,
25 doi:10.1007/BF00889358. <http://link.springer.com/10.1007/BF00889358> (Accessed November
26 4, 2018).
- 27 McIntosh, C., A. Sarris, and F. Papadopoulos, 2013: Productivity, credit, risk, and the demand for
28 weather index insurance in smallholder agriculture in Ethiopia. *Agric. Econ. (United Kingdom)*,
29 **44**, 399–417, doi:10.1111/agec.12024.
- 30 McKinley, D. C., and Coauthors, 2017: Citizen science can improve conservation science, natural
31 resource management, and environmental protection. *Biol. Conserv.*, **208**, 15–28,
32 doi:10.1016/j.biocon.2016.05.015.
- 33 McKinnon, A., 2010: Green logistics: the carbon agenda. *Electron. Sci. J. Logist.*, **6**.
- 34 McLeman, R. A., 2013: Climate and Human Migration. *Clim. Hum. Migr. Past Exp. Futur.*
35 *Challenges*, doi:10.1017/CBO9781139136938.
- 36 McMichael, A. J., R. E. Woodruff, and S. Hales, 2006: Climate change and human health: present and
37 future risks. *Lancet*, doi:10.1016/S0140-6736(06)68079-3.
- 38 McNicol, I. M., C. M. Ryan, and E. T. A. Mitchard, 2018a: Carbon losses from deforestation and
39 widespread degradation offset by extensive growth in African woodlands. *Nat. Commun.*, **9**,
40 3045, doi:10.1038/s41467-018-05386-z. <http://www.nature.com/articles/s41467-018-05386-z>
41 (Accessed April 14, 2019).
- 42 —, —, and —, 2018b: Carbon losses from deforestation and widespread degradation offset by
43 extensive growth in African woodlands. *Nat. Commun.*, **9**, 3045, doi:10.1038/s41467-018-
44 05386-z.
- 45 McSweeney, K., and O. T. Coomes, 2011: Climate-related disaster opens a window of opportunity for
46 rural poor in northeastern Honduras. *Proc. Natl. Acad. Sci.*, **108**, 5203–5208,
47 doi:10.1073/pnas.1014123108.

- 1 Measham, T. G., B. L. Preston, T. F. Smith, C. Brooke, R. Gorddard, G. Withycombe, and C.
2 Morrison, 2011: Adapting to climate change through local municipal planning: Barriers and
3 challenges. *Mitig. Adapt. Strateg. Glob. Chang.*, doi:10.1007/s11027-011-9301-2.
- 4 Mechler, R., L. M. Bouwer, J. Linnerooth-Bayer, S. Hochrainer-Stigler, J. C. J. H. Aerts, S.
5 Surminski, and K. Williges, 2014: Managing unnatural disaster risk from climate extremes. *Nat.*
6 *Clim. Chang.* 2014 44,.
- 7 Mehta, L., G. J. Veldwisch, and J. Franco, 2012: Introduction to the Special Issue: Water grabbing?
8 Focus on the (re)appropriation of finite water resources. *Water Altern.*, **5**, 193–207. www.water-
9 alternatives.org (Accessed April 13, 2019).
- 10 Meijer, S. S., D. Catacutan, O. C. Ajayi, G. W. Sileshi, and M. Nieuwenhuis, 2015: The role of
11 knowledge, attitudes and perceptions in the uptake of agricultural and agroforestry innovations
12 among smallholder farmers in sub-Saharan Africa. *Int. J. Agric. Sustain.*,
13 doi:10.1080/14735903.2014.912493.
- 14 van Meijl, H., and Coauthors, 2018a: Comparing impacts of climate change and mitigation on global
15 agriculture by 2050. *Environ. Res. Lett.*, **13**, 064021, doi:10.1088/1748-9326/aabdc4.
16 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=6/a=064021?key=crossref.42a4eb1897f2ed545f2b0dc439d03e64)
17 [9326/13/i=6/a=064021?key=crossref.42a4eb1897f2ed545f2b0dc439d03e64](http://stacks.iop.org/1748-9326/13/i=6/a=064021?key=crossref.42a4eb1897f2ed545f2b0dc439d03e64) (Accessed April 12,
18 2019).
- 19 —, and Coauthors, 2018b: Comparing impacts of climate change and mitigation on global
20 agriculture by 2050. *Environ. Res. Lett.*, **13**, 064021, doi:10.1088/1748-9326/aabdc4.
- 21 Meinzen-dick, R., A. Quisumbing, J. Behrman, P. Biermayr-Jenzano, V. Wilde, M. Noordeloos, C.
22 Ragasa, and N. Beintema, 2010: Engendering Agricultural Research. *IFPRI Discussion Pap.*
23 *973*, 72.
- 24 Mekonnen, M. M., and A. Y. Hoekstra, 2016: Sustainability: Four billion people facing severe water
25 scarcity. *Sci. Adv.*, doi:10.1126/sciadv.1500323.
- 26 Meli, P., K. D. Holl, J. M. R. Benayas, H. P. Jones, P. C. Jones, D. Montoya, and D. M. Mateos,
27 2017: A global review of past land use, climate, and active vs. passive restoration effects on
28 forest recovery. *PLoS One*, **12**, e0171368.
- 29 Mello, D., and M. Schmink, 2017a: Amazon entrepreneurs: Women’s economic empowerment and
30 the potential for more sustainable land use practices. *Womens. Stud. Int. Forum*, **65**, 28–36,
31 doi:10.1016/J.WSIF.2016.11.008.
32 <https://www.sciencedirect.com/science/article/abs/pii/S027753951530176X> (Accessed April 2,
33 2019).
- 34 —, and —, 2017b: Amazon entrepreneurs: Women’s economic empowerment and the potential
35 for more sustainable land use practices. *Womens. Stud. Int. Forum*, **65**, 28–36,
36 doi:10.1016/J.WSIF.2016.11.008.
- 37 Mello, F. F. C., and Coauthors, 2014: Payback time for soil carbon and sugar-cane ethanol. *Nat. Clim.*
38 *Chang.*, doi:10.1038/nclimate2239.
- 39 Melvin, A. M., and Coauthors, 2017: Climate change damages to Alaska public infrastructure and the
40 economics of proactive adaptation. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1611056113.
- 41 Menz, M. H. M., K. W. Dixon, and R. J. Hobbs, 2013: Hurdles and opportunities for landscape-scale
42 restoration. *Science (80-.)*, **339**, 526–527.
- 43 Mersha, A. A., and F. Van Laerhoven, 2016: A gender approach to understanding the differentiated
44 impact of barriers to adaptation: responses to climate change in rural Ethiopia. *Reg. Environ.*
45 *Chang.*, **16**, 1701–1713, doi:10.1007/s10113-015-0921-z.
- 46 Messerli, P., M. Giger, M. B. Dwyer, T. Breu, and S. Eckert, 2014a: The geography of large-scale
47 land acquisitions: Analysing socio-ecological patterns of target contexts in the global South.
48 *Appl. Geogr.*, **53**, 449–459, doi:10.1016/J.APGEOG.2014.07.005.

- 1 <https://www.sciencedirect.com/science/article/pii/S0143622814001611> (Accessed April 9,
2 2019).
- 3 —, —, —, —, and —, 2014b: The geography of large-scale land acquisitions: Analysing
4 socio-ecological patterns of target contexts in the global South. *Appl. Geogr.*, **53**, 449–459,
5 doi:10.1016/J.APGEOG.2014.07.005.
- 6 Metternicht, G., 2018: Contributions of Land Use Planning to Sustainable Land Use and
7 Management. Springer, Cham, 35–51 http://link.springer.com/10.1007/978-3-319-71861-3_4
8 (Accessed October 31, 2018).
- 9 Metz, B., J. K. Turkson, and Intergovernmental Panel on Climate Change. Working Group III., 2000:
10 *Methodological and technological issues in technology transfer*. Published for the
11 Intergovernmental Panel on Climate Change [by] Cambridge University Press, 466 pp.
- 12 Meyer, M. A., and J. A. Priess, 2014: Indicators of bioenergy-related certification schemes – An
13 analysis of the quality and comprehensiveness for assessing local/regional environmental
14 impacts. *Biomass and Bioenergy*, **65**, 151–169, doi:10.1016/J.BIOMBIOE.2014.03.041.
15 <https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0961953414001706> (Accessed
16 May 23, 2018).
- 17 Meyfroidt, P., E. F. Lambin, K. H. Erb, and T. W. Hertel, 2013: Globalization of land use: Distant
18 drivers of land change and geographic displacement of land use. *Curr. Opin. Environ. Sustain.*,
19 doi:10.1016/j.cosust.2013.04.003.
- 20 Michaelowa, K., and A. Michaelowa, 2017: Transnational Climate Governance Initiatives: Designed
21 for Effective Climate Change Mitigation? *Int. Interact.*, **43**, 129–155,
22 doi:10.1080/03050629.2017.1256110.
23 <https://www.tandfonline.com/doi/full/10.1080/03050629.2017.1256110> (Accessed December
24 12, 2017).
- 25 Michel-Kerjan, E., 2011: *Catastrophe Financing for Governments: Learning from the 2009-2012*
26 *MultiCat Program in Mexico*.
- 27 Middleton, N., U. Kang, N. Middleton, and U. Kang, 2017: Sand and Dust Storms: Impact Mitigation.
28 *Sustainability*, **9**, 1053, doi:10.3390/su9061053.
- 29 Milder, J. C., and Coauthors, 2015: An agenda for assessing and improving conservation impacts of
30 sustainability standards in tropical agriculture. *Conserv. Biol.*, **29**, 309–320,
31 doi:10.1111/cobi.12411. <http://doi.wiley.com/10.1111/cobi.12411> (Accessed April 3, 2019).
- 32 Milkoreit, M., M. L. Moore, M. Schoon, and C. L. Meek, 2015: Resilience scientists as change-
33 makers-Growing the middle ground between science and advocacy? *Environ. Sci. Policy*,
34 doi:10.1016/j.envsci.2014.08.003.
- 35 Millar, C. I., N. L. Stephenson, and S. L. Stephens, 2007: Climate change and forests of the future:
36 managing in the face of uncertainty. *Ecol. Appl.*, **17**, 2145–2151.
- 37 Miller, L., R. Carriveau, and S. Harper, 2018: Innovative financing for renewable energy project
38 development – recent case studies in North America. *Int. J. Environ. Stud.*, **75**, 121–134,
39 doi:10.1080/00207233.2017.1403758.
40 <https://www.tandfonline.com/doi/full/10.1080/00207233.2017.1403758>.
- 41 Millo, G., 2016: The Income Elasticity of Nonlife Insurance: A Reassessment. *J. Risk Insur.*, **83**, 335–
42 362, doi:10.1111/jori.12051. <http://doi.wiley.com/10.1111/jori.12051> (Accessed April 3, 2019).
- 43 Milton, S. J., W. R. J. Dean, and D. M. Richardson, 2003: Economic incentives for restoring natural
44 capital in southern African rangelands. *Front. Ecol. Environ.*, **1**, 247–254.
- 45 Mimura, N., R. S. Pulwarty, D. M. Duc, I. Elshinnawy, M. H. Redsteer, H. Q. Huang, J. N. Nkem,
46 and R. a. S. Rodriguez, 2014: 15. Adaptation Planning and Implementation. *Assess. Rep. 5-*
47 *Clim. Chang. 2014 Impacts, Adapt. Vulnerability. Part A Glob. Sect. Asp.*, 869–898,
48 doi:10.1029/2003JD004173.Aires.

- 1 Minang, P. A., and Coauthors, 2014: REDD+ Readiness progress across countries: time for
2 reconsideration. *Clim. Policy*, **14**, 685–708, doi:10.1080/14693062.2014.905822.
3 <http://www.tandfonline.com/doi/abs/10.1080/14693062.2014.905822> (Accessed October 15,
4 2018).
- 5 Mini, C., T. S. Hogue, and S. Pincetl, 2015: The effectiveness of water conservation measures on
6 summer residential water use in Los Angeles, California. *Resour. Conserv. Recycl.*,
7 doi:10.1016/j.resconrec.2014.10.005.
- 8 Ministry of Environment and Forests and Ministry, and N. D. Tribal Affairs, Government of India,
9 2010: *Manthan: Report of the National Committee on Forest Rights Act*.
- 10 Mistry, J., and A. Berardi, 2016: Bridging indigenous and scientific knowledge. *Science (80-.)*,
11 doi:10.1126/science.aaf1160.
- 12 Mitchell, D., 2010: Land Tenure and Disaster Risk Management. *L. Tenure J.*, **1**, 121–141.
13 <http://www.fao.org/nr/tenure/land-tenure-journal/index.php/LTJ/article/viewArticle/11/40>
14 (Accessed November 3, 2018).
- 15 Mitchell, D., S. Enemark, and P. van der Molen, 2015: Climate resilient urban development: Why
16 responsible land governance is important. *Land use policy*, **48**, 190–198,
17 doi:10.1016/J.LANDUSEPOL.2015.05.026.
18 <https://www.sciencedirect.com/science/article/abs/pii/S0264837715001660> (Accessed
19 November 3, 2018).
- 20 Miteva, D. A., C. J. Loucks, and S. K. Pattanayak, 2015: Social and Environmental Impacts of Forest
21 Management Certification in Indonesia. *PLoS One*, **10**, e0129675,
22 doi:10.1371/journal.pone.0129675. <https://dx.plos.org/10.1371/journal.pone.0129675> (Accessed
23 April 8, 2019).
- 24 Mobarak, A. M., and M. R. Rosenzweig, 2013: Informal risk sharing, index insurance, and risk taking
25 in developing countries. *American Economic Review*, Vol. 103 of, 375–380.
- 26 Mochizuki, J., S. Vitoontus, B. Wickramarachchi, S. Hochrainer-Stigler, K. Williges, R. Mechler, and
27 R. Sovann, 2015: Operationalizing iterative risk management under limited information: Fiscal
28 and economic risks due to natural disasters in Cambodia. *Int. J. Disaster Risk Sci.*, **6**, 321–334,
29 doi:10.1007/s13753-015-0069-y.
- 30 Mohapatra, S., 2013: Displacement due to Climate Change and International Law. *Int. J. Manag. Soc.*
31 *Sci. Res.*,
- 32 Mohammed, A., and Coauthors, 2018: Assessing drought vulnerability and adaptation among farmers
33 in Gadaref region, Eastern Sudan. *Land use policy*, **70**, 402–413,
34 doi:10.1016/j.landusepol.2017.11.027.
35 <http://linkinghub.elsevier.com/retrieve/pii/S0264837717305355> (Accessed May 16, 2018).
- 36 Mohr, A., T. Beuchelt, R. € El Schneider, and D. Virchow, 2016: Food security criteria for voluntary
37 biomass sustainability standards and certifications. *Biomass and Bioenergy*, **89**, 133–145,
38 doi:10.1016/j.biombioe.2016.02.019. <http://dx.doi.org/10.1016/j.biombioe.2016.02.019>
39 (Accessed April 13, 2019).
- 40 van der Molen, P., and D. Mitchell, 2016: Climate change, land use and land surveyors. *Surv. Rev.*,
41 **48**, 148–155, doi:10.1179/1752270615Y.0000000029.
42 <https://www.tandfonline.com/doi/full/10.1179/1752270615Y.0000000029> (Accessed May 22,
43 2018).
- 44 Mollenkamp, S., and B. Kasten, 2009: Institutional Adaptation to Climate Change: The Current Status
45 and Future Strategies in the Elbe Basin, Germany. *Climate change adaptation in the water*
46 *sector*, 227–249.
- 47 Monchuk, V., 2014: Reducing Poverty and Investing in People The New Role of Safety Nets in
48 Africa Human Development. *World Bank*,

- 1 <https://openknowledge.worldbank.org/bitstream/handle/10986/16256/9781464800948.pdf>
2 (Accessed May 26, 2018).
- 3 Monkelbaan, J., 2019: *Governance for the Sustainable Development Goals : exploring an integrative*
4 *framework of theories, tools, and competencies*. 1st ed. Springer Singapore, XXI, 214 pp.
- 5 Montaña, E., H. P. Diaz, and M. Hurlbert, 2016: Development, local livelihoods, and vulnerabilities
6 to global environmental change in the South American Dry Andes. *Reg. Environ. Chang.*,
7 doi:10.1007/s10113-015-0888-9.
- 8 Montanarella, L., 2015: The Importance of Land Restoration for Achieving a Land Degradation-
9 Neutral World. *Land Restoration: Reclaiming Landscapes for a Sustainable Future*, 249–258.
- 10 Monterosso, I., P. Cronkleton, D. Pinedo, and A. . Larson, 2017: *Reclaiming collective rights: land*
11 *and forest tenure reforms in Peru (1960-2016)*. Center for International Forestry Research
12 (CIFOR), [http://www.cifor.org/library/6426/reclaiming-collective-rights-land-and-forest-tenure-](http://www.cifor.org/library/6426/reclaiming-collective-rights-land-and-forest-tenure-reforms-in-peru-1960-2016/)
13 [reforms-in-peru-1960-2016/](http://www.cifor.org/library/6426/reclaiming-collective-rights-land-and-forest-tenure-reforms-in-peru-1960-2016/) (Accessed April 9, 2019).
- 14 Moore, C. T., E. V. Lonsdorf, M. G. Knutson, H. P. Laskowski, and S. K. Lor, 2011: Adaptive
15 management in the U.S. National Wildlife Refuge System: Science-management partnerships
16 for conservation delivery. *J. Environ. Manage.*, **92**, 1395–1402,
17 doi:10.1016/j.jenvman.2010.10.065.
- 18 Moore, F. C., and D. B. Diaz, 2015: Temperature impacts on economic growth warrant stringent
19 mitigation policy. **5**, 127–132, doi:10.1038/NCLIMATE2481.
- 20 Moosa, C. S., and N. Tuana, 2014: Mapping a Research Agenda Concerning Gender and Climate
21 Change: A Review of the Literature. *Hypatia*, **29**, 677–694, doi:10.1111/hypa.12085.
22 <http://doi.wiley.com/10.1111/hypa.12085> (Accessed April 2, 2019).
- 23 Moran, D., 2011: Marginal Abatement Cost Curves for UK Agricultural Greenhouse Gas Emissions.
24 *J. Agric. Econ.*, **62**, 93–118.
- 25 Morita, K., and K. Matsumoto, 2018: Synergies among climate change and biodiversity conservation
26 measures and policies in the forest sector: A case study of Southeast Asian countries. *For.*
27 *Policy Econ.*, **87**, 59–69, doi:https://doi.org/10.1016/j.forpol.2017.10.013.
28 <http://www.sciencedirect.com/science/article/pii/S1389934117302009>.
- 29 Moroni, S., 2018: Property as a human right and property as a special title. Rediscussing private
30 ownership of land. *Land use policy*, **70**, 273–280, doi:10.1016/J.LANDUSEPOL.2017.10.037.
31 <https://www.sciencedirect.com/science/article/pii/S0264837717303368?via%3Dihub> (Accessed
32 May 23, 2018).
- 33 Morton, J., 2017: Climate change and African agriculture: unlocking the potential of research and
34 advisory services. *Making Climate Compatible Development Happen*, F. Nunan, Ed., Routledge,
35 87–113 <http://gala.gre.ac.uk/16696/> (Accessed November 2, 2018).
- 36 Morton, J. F., 2007: The impact of climate change on smallholder and subsistence agriculture. *Proc.*
37 *Natl. Acad. Sci. U. S. A.*, **104**, 19680–19685, doi:10.1073/pnas.0701855104.
38 <http://www.ncbi.nlm.nih.gov/pubmed/18077400> (Accessed May 22, 2018).
- 39 Mostert, E., C. Pahl-Wostl, Y. Rees, B. Searle, D. Tàbara, and J. Tippett, 2007: Social learning in
40 European river-basin management: Barriers and fostering mechanisms from 10 river basins.
41 *Ecol. Soc.*, **12**, doi:19.
- 42 Mowo, J., Z. Adimassu, D. Catacutan, J. Tanui, K. Masuki, and C. Lyamchai, 2013: The Importance
43 of Local Traditional Institutions in the Management of Natural Resources in the Highlands of
44 East Africa. *Hum. Organ.*, **72**, 154–163, doi:10.17730/humo.72.2.e1x3101741127x35.
- 45 Mozumder, P., R. Helton, and R. P. Berrens, 2009: Provision of a wildfire risk map: informing
46 residents in the wildland urban interface. *Risk Anal. An Int. J.*, **29**, 1588–1600.
- 47 —, E. Flugman, and T. Randhir, 2011: Adaptation behavior in the face of global climate change:

- 1 Survey responses from experts and decision makers serving the Florida Keys. *Ocean Coast.*
2 *Manag.*, doi:10.1016/j.ocecoaman.2010.10.008.
- 3 Mubaya, C. P., and P. Mafongoya, 2017: The role of institutions in managing local level climate
4 change adaptation in semi-arid Zimbabwe. *Clim. Risk Manag.*, doi:10.1016/j.crm.2017.03.003.
- 5 Mudombi, S., and Coauthors, 2018a: Multi-dimensional poverty effects around operational biofuel
6 projects in Malawi, Mozambique and Swaziland. *Biomass and Bioenergy*, **114**, 41–54,
7 doi:10.1016/j.biombioe.2016.09.003.
8 <https://linkinghub.elsevier.com/retrieve/pii/S0961953416302938> (Accessed April 13, 2019).
- 9 —, and Coauthors, 2018b: Multi-dimensional poverty effects around operational biofuel projects in
10 Malawi, Mozambique and Swaziland. *Biomass and Bioenergy*, **114**, 41–54,
11 doi:10.1016/j.biombioe.2016.09.003.
- 12 —, A. Nyambane, G. P. von Maltitz, A. Gasparatos, F. X. Johnson, M. L. Chenene, and B.
13 Attanassov, 2018c: User perceptions about the adoption and use of ethanol fuel and cookstoves
14 in Maputo, Mozambique. *Energy Sustain. Dev.*, **44**, 97–108, doi:10.1016/j.esd.2018.03.004.
15 <https://linkinghub.elsevier.com/retrieve/pii/S0973082617305458> (Accessed April 13, 2019).
- 16 Mukherjee, N., J. Hugé, W. J. Sutherland, J. McNeill, M. Van Opstal, F. Dahdouh-Guebas, and N.
17 Koedam, 2015: The Delphi technique in ecology and biological conservation: Applications and
18 guidelines. *Methods Ecol. Evol.*, **6**, 1097–1109, doi:10.1111/2041-210X.12387.
- 19 Mukhtarov, F., P. Osseweijer, and R. Pierce, Global governance of biofuels: a case for public-private
20 governance? doi:10.13128/BAE-14767. www.fupress.com/bae (Accessed April 12, 2019a).
- 21 —, —, and —, Global governance of biofuels: a case for public-private governance?
22 doi:10.13128/BAE-14767.
- 23 Müller, A., J. Weigelt, A. Götz, O. Schmidt, I. Lobos Alva, I. Matuschke, U. Ehling, and T. Beringer,
24 2015: *IASS Working Paper The Role of Biomass in the Sustainable Development Goals: A*
25 *Reality Check and Governance Implications*. [http://publications.iass-](http://publications.iass-potsdam.de/pubman/item/escidoc:1014893:4/component/escidoc:1014896/IASS_Working_Paper_1014893.pdf)
26 [potsdam.de/pubman/item/escidoc:1014893:4/component/escidoc:1014896/IASS_Working_Pape](http://publications.iass-potsdam.de/pubman/item/escidoc:1014893:4/component/escidoc:1014896/IASS_Working_Paper_1014893.pdf)
27 [r_1014893.pdf](http://publications.iass-potsdam.de/pubman/item/escidoc:1014893:4/component/escidoc:1014896/IASS_Working_Paper_1014893.pdf) (Accessed April 12, 2019).
- 28 Muller, C., and H. Yan, 2018: Household fuel use in developing countries: Review of theory and
29 evidence. *Energy Econ.*, **70**, 429–439, doi:10.1016/j.eneco.2018.01.024. [https://www-](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S014098831830032X)
30 [sciencedirect-com.ezp.sub.su.se/science/article/pii/S014098831830032X](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S014098831830032X) (Accessed April 14,
31 2019).
- 32 Müller, C., and Coauthors, 2013: Assessing agricultural risks of climate change in the 21st century in
33 a global gridded crop model intercomparison. *Proc. Natl. Acad. Sci.*,
34 doi:10.1073/pnas.1222463110.
- 35 Munang, R., I. Thiaw, K. Alverson, M. Mumba, J. Liu, and M. Rivington, 2013: Climate change and
36 Ecosystem-based Adaptation: A new pragmatic approach to buffering climate change impacts.
37 *Curr. Opin. Environ. Sustain.*, **5**, 67–71, doi:10.1016/j.cosust.2012.12.001.
- 38 Munene, M. B., Å. G. Swartling, and F. Thomalla, 2018: Adaptive governance as a catalyst for
39 transforming the relationship between development and disaster risk through the Sendai
40 Framework? *Int. J. Disaster Risk Reduct.*, doi:10.1016/j.ijdr.2018.01.021.
- 41 Munthali, K., and Y. Murayama, 2013: Interdependences between Smallholder Farming and
42 Environmental Management in Rural Malawi: A Case of Agriculture-Induced Environmental
43 Degradation in Malingunde Extension Planning Area (EPA). *Land*, doi:10.3390/land2020158.
- 44 Muradian, R., and L. Rival, 2012: Between markets and hierarchies: the challenge of governing
45 ecosystem services. *Ecosyst. Serv.*, **1**, 93–100.
- 46 Muratori, M., K. Calvin, M. Wise, P. Kyle, and J. Edmonds, 2016: Global economic consequences of
47 deploying bioenergy with carbon capture and storage (BECCS). *Environ. Res. Lett.*,
48 doi:10.1088/1748-9326/11/9/095004.

- 1 Murphy, K., G. A. Kirkman, S. Seres, and E. Haites, 2015: Technology transfer in the CDM: an
2 updated analysis. *Clim. Policy*, **15**, 127–145, doi:10.1080/14693062.2013.812719.
3 <http://www.tandfonline.com/doi/abs/10.1080/14693062.2013.812719> (Accessed May 22, 2018).
- 4 Murthy, I. K., V. Varghese, P. Kumar, and S. Sridhar, 2018a: Experience of Participatory Forest
5 Management in India: Lessons for Governance and Institutional Arrangements Under REDD+.
6 *Global Forest Governance and Climate Change*, Springer International Publishing, Cham, 175–
7 201 http://link.springer.com/10.1007/978-3-319-71946-7_7 (Accessed March 9, 2019).
- 8 —, —, —, and —, 2018b: Experience of Participatory Forest Management in India: Lessons
9 for Governance and Institutional Arrangements Under REDD+. *Global Forest Governance and*
10 *Climate Change*, Springer International Publishing, Cham, 175–201.
- 11 Mustalahti, I., and O. S. Rakotonarivo, 2014: REDD+ and Empowered Deliberative Democracy:
12 Learning from Tanzania. *World Dev.*, **59**, 199–211, doi:10.1016/j.worlddev.2014.01.022.
- 13 Naeem, S., and Coauthors, 2015: Get the science right when paying for nature’s services. *Science* (80-
14 .), **347**, 1206–1207.
- 15 Naess, L. O., 2013: The role of local knowledge in adaptation to climate change. *Wiley Interdiscip.*
16 *Rev. Chang.*, doi:Doi 10.1002/Wcc.204.
- 17 Nagel, L. M., and Coauthors, 2017: Adaptive silviculture for climate change: a national experiment in
18 manager-scientist partnerships to apply an adaptation framework. *J. For.*, **115**, 167–178.
- 19 Nagendra, H., and E. Ostrom, 2012: Polycentric governance of multifunctional forested landscapes.
20 *Int. J. Commons*, **6**, 104–133, doi:10.18352/ijc.321.
- 21 Naicker, P. ., 2011: The impact of climate change and other factors on zoonotic diseases. *Arch. Clin.*
22 *Microbiol.*, **2**.
23 https://www.researchgate.net/publication/287523149_The_impact_of_climate_change_and_othe
24 [r_factors_on_zoonotic_diseases](https://www.researchgate.net/publication/287523149_The_impact_of_climate_change_and_othe) (Accessed April 25, 2019).
- 25 Nakashima, D., K. G. McLean, H. D. Thulstrup, A. R. Castillo, and J. T. Rubis, 2013: *Weathering*
26 *uncertainty: traditional knowledge for climate change assessment and adaptation*. D.
27 McDonald, Ed. UNESCO and UNU, Paris, 120 pp. www.unutki.org (Accessed April 15, 2019).
- 28 Nakhooda, S., C. Watson, and L. Schalatek, 2016: The Global Climate Finance Architecture. *Clim.*
29 *Financ. Fundam.*, **5**.
- 30 Nalau, J., and J. Handmer, 2015: When is transformation a viable policy alternative? *Environ. Sci.*
31 *Policy*, doi:10.1016/j.envsci.2015.07.022.
- 32 Namubiru-Mwaura, E., 2014a: *LAND TENURE AND GENDER: APPROACHES AND*
33 *CHALLENGES FOR STRENGTHENING RURAL WOMEN’S LAND RIGHTS*.
34 www.worldbank.org/gender/agency (Accessed April 2, 2019).
- 35 —, 2014b: *LAND TENURE AND GENDER: APPROACHES AND CHALLENGES FOR*
36 *STRENGTHENING RURAL WOMEN’S LAND RIGHTS*.
- 37 Nanda, A. V. ., J. Rijke, L. Beesley, B. Gersonius, M. R. Hipsey, and A. Ghadouani, 2018: Matching
38 Ecosystem Functions with Adaptive Ecosystem Management : Decision Pathways to Overcome
39 Institutional Barriers. *Water*, **10**, doi:10.3390/w10060672.
- 40 Narassimhan, E., K. S. Gallagher, S. Koester, J. Rivera, K. S. Gallagher, and S. Koester, 2018:
41 Carbon pricing in practice: a review of existing emissions trading systems. **3062**,
42 doi:10.1080/14693062.2018.1467827.
- 43 Nawrotzki, R. J., and M. Bakhtsiyarava, 2017: International Climate Migration: Evidence for the
44 Climate Inhibitor Mechanism and the Agricultural Pathway. *Popul. Space Place*,
45 doi:10.1002/psp.2033.
- 46 —, A. M. Schlak, and T. A. Kugler, 2016: Climate, migration, and the local food security context:
47 introducing Terra Populus. *Popul. Environ.*, doi:10.1007/s11111-016-0260-0.

- 1 Nayak, R. R., S. Vaidyanathan, and J. Krishnaswamy, 2014: Fire and grazing modify grass
2 community response to environmental determinants in savannas: Implications for sustainable
3 use. *Agric. Ecosyst. Environ.*, **185**, 197–207.
- 4 Ndegwa, G. M., U. Nehren, F. Grüniger, M. Iiyama, and D. Anhof, 2016: Charcoal production
5 through selective logging leads to degradation of dry woodlands: a case study from Mutomo
6 District, Kenya. *J. Arid Land*, **8**, 618–631, doi:10.1007/s40333-016-0124-6.
7 <http://link.springer.com/10.1007/s40333-016-0124-6> (Accessed May 23, 2018).
- 8 Nelson, G. C., V. Harris, and S. W. Stone, 2001: Deforestation, Land Use, and Property Rights:
9 Empirical Evidence from Darien, Panama. *Land Econ.*, **77**, 187, doi:10.2307/3147089.
- 10 Nelson, R., M. Howden, and M. S. Smith, 2008: Using adaptive governance to rethink the way
11 science supports Australian drought policy. *Environ. Sci. Policy*, **11**, 588–601,
12 doi:10.1016/j.envsci.2008.06.005.
- 13 Nelson, R., P. Kokic, S. Crimp, P. Martin, H. Meinke, S. M. Howden, P. de Voil, and U. Nidumolu,
14 2010: The vulnerability of Australian rural communities to climate variability and change: Part
15 II-Integrating impacts with adaptive capacity. *Environ. Sci. Policy*, **13**, 18–27,
16 doi:10.1016/j.envsci.2009.09.007.
- 17 NEPSTAD, D., and Coauthors, 2006: Inhibition of Amazon Deforestation and Fire by Parks and
18 Indigenous Lands. *Conserv. Biol.*, **20**, 65–73, doi:10.1111/j.1523-1739.2006.00351.x.
19 <http://doi.wiley.com/10.1111/j.1523-1739.2006.00351.x> (Accessed April 9, 2019).
- 20 Nepstad, D. C., C. M. Stickler, B. Soares-Filho, and F. Merry, 2008: Interactions among Amazon land
21 use, forests and climate: prospects for a near-term forest tipping point. *Philos. Trans. R. Soc.
22 London B Biol. Sci.*, **363**, 1737–1746.
- 23 Neufeldt, H., and Coauthors, 2013: Beyond climate-smart agriculture: toward safe operating spaces
24 for global food systems. *Agric. Food Secur.*, **2**, (30 August 2013), doi:10.1186/2048-7010-2-12.
- 25 New Zealand Ministry for the Environment, 2018: *New Zealand's Greenhouse Gas Inventory 1990-*
26 *2016*. Wellington, [https://www.mfe.govt.nz/sites/default/files/media/Climate Change/National
27 GHG Inventory Report 1990-2016-final.pdf](https://www.mfe.govt.nz/sites/default/files/media/Climate%20Change/National%20GHG%20Inventory%20Report%201990-2016-final.pdf).
- 28 New Zealand Productivity Commission, 2018: *Low-Emissions Economy: Final Report*. Wellington,
29 [https://www.productivity.govt.nz/sites/default/files/Productivity
30 Commission_Low-emissions
economy_Final Report_FINAL_2.pdf](https://www.productivity.govt.nz/sites/default/files/Productivity%20Commission_Low-emissions%20economy_Final%20Report_FINAL_2.pdf).
- 31 Newell, P., and O. Taylor, 2017: Contested landscapes: the global political economy of climate-smart
32 agriculture. *J. Peasant Stud.*, **0**, 1–22, doi:10.1080/03066150.2017.1324426.
- 33 Newig, J., D. Gunther, and C. Pahl-Wostl, 2010: Synapses in the Network : Learning in Governance
34 Networks in the. *Ecol. Soc.*, **15**, 24, doi:10.1197/jamia.M2338.
- 35 Newton, A. C., 2016: Biodiversity Risks of Adopting Resilience as a Policy Goal. *Conserv. Lett.*,
36 doi:10.1111/conl.12227.
- 37 Newton, P., R. Benzeev, A. Agrawal, C. Liao, C. Watkins, L. V. Rasmussen, and R. Hajjar, 2018: The
38 role of zero-deforestation commitments in protecting and enhancing rural livelihoods. *Curr.
39 Opin. Environ. Sustain.*, **32**, 126–133, doi:10.1016/j.cosust.2018.05.023.
40 <https://doi.org/10.1016/j.cosust.2018.05.023> (Accessed April 7, 2019).
- 41 Nguyen, T., and J. Lindenmeier, 2014: Catastrophe risks, cat bonds and innovation resistance. *Qual.
42 Res. Financ. Mark.*, **6**, 75–92, doi:10.1108/QRFM-06-2012-0020.
- 43 Nicholls, R. J., and A. Cazenave, 2010: Sea-level rise and its impact on coastal zones. *Science (80-.)*,
44 **328**, 1517–1520.
- 45 Nieto-Romero, M., A. Milcu, J. Leventon, F. Mikulcak, and J. Fischer, 2016: The role of scenarios in
46 fostering collective action for sustainable development: Lessons from central Romania. *Land use
47 policy*, **50**, 156–168, doi:10.1016/J.LANDUSEPOL.2015.09.013.

- 1 <https://www.sciencedirect.com/science/article/pii/S0264837715002847> (Accessed May 23,
2 2018).
- 3 Nikolakis, W., S. Akter, and H. Nelson, 2016: The effect of communication on individual preferences
4 for common property resources: A case study of two Canadian First Nations. *Land use policy*,
5 **58**, 70–82, doi:10.1016/J.LANDUSEPOL.2016.07.007.
6 <https://www.sciencedirect.com/science/article/pii/S0264837716306913> (Accessed May 23,
7 2018).
- 8 Nilsson, C., and K. Berggren, 2000: Alterations of Riparian Ecosystems Caused by River Regulation:
9 Dam operations have caused global-scale ecological changes in riparian ecosystems. How to
10 protect river environments and human needs of rivers remains one of the most important
11 questions of ou. *AIBS Bull.*, **50**, 783–792.
- 12 Nilsson, M., and Å. Persson, 2012: Can Earth system interactions be governed? Governance functions
13 for linking climate change mitigation with land use, freshwater and biodiversity protection. *Ecol.*
14 *Econ.*, **75**, 61–71, doi:10.1016/J.ECOLECON.2011.12.015.
15 <http://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0921800911005313> (Accessed
16 December 12, 2017).
- 17 Nilsson, M., D. Griggs, and M. Visback, 2016a: Map the interactions between Sustainable
18 Development Goa. *Nature*, **534**, 320–322, doi:10.1038/534320a. <http://dx.doi.org/10.1787/>.
- 19 Nilsson, M., D. Griggs, and M. Visbeck, 2016b: Map the interactions between sustainable
20 development goals: Mans Nilsson, Dave Griggs and Martin Visbeck present a simple way of
21 rating relationships between the targets to highlight priorities for integrated policy. *Nature*, **534**,
22 320–323.
- 23 Niño-Zarazúa, M., A. Barrientos, S. Hickey, and D. Hulme, 2012: Social Protection in Sub-Saharan
24 Africa: Getting the Politics Right. *World Dev.*, doi:10.1016/j.worlddev.2011.04.004.
- 25 Nischalke, S. M., 2015: Adaptation Options Adaptation options to Improve Food Security in a
26 Changing Climate in the Hindu Kush-Himalayan Region. *Handbook of Climate Change*
27 *Adaptation*, Springer Berlin Heidelberg, Berlin, Heidelberg, 1423–1442
28 http://link.springer.com/10.1007/978-3-642-38670-1_103 (Accessed May 16, 2018).
- 29 Njenga, M., and R. Mendum, 2018: *Recovering bioenergy in Sub-Saharan Africa: gender dimensions,*
30 *lessons and challenges.* Nairobi, Kenya, 96 pp.
31 <http://www.iwmi.cgiar.org/publications/resource-recovery-reuse/special-issue/> (Accessed April
32 14, 2019).
- 33 Nkengasong, J. N., and P. Onyebujoh, 2018: Response to the Ebola virus disease outbreak in the
34 Democratic Republic of the Congo. *Lancet*, **391**, 2395–2398, doi:10.1016/S0140-
35 6736(18)31326-6.
- 36 Nkoana, E. M., T. Waas, A. Verbruggen, C. J. Burman, and J. Hugé, 2017: Analytic framework for
37 assessing participation processes and outcomes of climate change adaptation tools. *Environ.*
38 *Dev. Sustain.*, **19**, 1731–1760, doi:10.1007/s10668-016-9825-4.
39 <http://link.springer.com/10.1007/s10668-016-9825-4> (Accessed May 23, 2018).
- 40 Nkonya, E., J. von Braun, A. Mirzabaev, Q. B. Le, H. Y. Kwon, and O. Kirui, 2013: Economics of
41 Land Degradation Initiative: Methods and Approach for Global and National Assessments.
42 *SSRN Electron. J.*, doi:10.2139/ssrn.2343636. <http://www.ssrn.com/abstract=2343636> (Accessed
43 May 23, 2018).
- 44 —, T. Johnson, H. Y. Kwon, and E. Kato, 2015: Economics of land degradation in sub-Saharan
45 Africa. *Economics of Land Degradation and Improvement - A Global Assessment for*
46 *Sustainable Development*.
- 47 —, W. Anderson, E. Kato, J. Koo, A. Mirzabaev, J. von Braun, and S. Meyer, 2016: Global Cost of
48 Land Degradation BT - Economics of Land Degradation and Improvement – A Global
49 Assessment for Sustainable Development. E. Nkonya, A. Mirzabaev, and J. Von Braun, Eds.,

- 1 Springer International Publishing, Cham, 117–165 [https://doi.org/10.1007/978-3-319-19168-](https://doi.org/10.1007/978-3-319-19168-3_6)
2 3_6.
- 3 Noble, I. R., S. Huq, Y. a. Anokhin, J. Carmin, D. Goudou, F. P. Lansigan, B. Osman-Elasha, and A.
4 Villamizar, 2014: *IPCC 5th report - WGII: Adaptation Needs and Options*. 833-868 pp.
- 5 Nolte, K., W. Chamberlain, and M. Giger, 2016: *International Land Deals for Agriculture. Fresh*
6 *insights from the Land Matrix: Analytical Report II*. Bern, Montpellier, Hamburg, Pretoria,
7 https://www.researchgate.net/publication/308983402_International_Land_Deals_for_Agricultur
8 [e_Fresh_insights_from_the_Land_Matrix_Analytical_Report_II](https://www.researchgate.net/publication/308983402_International_Land_Deals_for_Agricultur).
- 9 Nordhaus, W., 2014: Estimates of the social cost of carbon: concepts and results from the DICE-
10 2013R model and alternative approaches. *J. Assoc. Environ. Resour. Econ.*, **1**, 273–312.
- 11 Nordhaus, W. D., 1999: Roll the DICE Again: The Economics of Global Warming. *Draft Version*, **28**,
12 1999.
- 13 Norström, A., and Coauthors, 2014: Three necessary conditions for establishing effective Sustainable
14 Development Goals in the Anthropocene. *Ecol. Soc.*, **19**.
- 15 North, D., 1991: Institutions. *J. Econ. Perspect.*, doi:10.1257/jep.5.1.97.
- 16 Northrup, J. M., and G. Wittemyer, 2013: Characterising the impacts of emerging energy
17 development on wildlife, with an eye towards mitigation. *Ecol. Lett.*, **16**, 112–125.
- 18 Nugent, R., M. Y. Bertram, S. Jan, L. W. Niessen, F. Sassi, D. T. Jamison, E. G. Pier, and R.
19 Beaglehole, 2018: Investing in non-communicable disease prevention and management to
20 advance the Sustainable Development Goals. *The Lancet*.
- 21 Nuhoff-Isakhanyan, G., E. Wubben, and S. W. F. Omta, 2016: Sustainability Benefits and Challenges
22 of Inter-Organizational Collaboration in Bio-Based Business: A Systematic Literature Review.
23 *Sustainability*, **8**, 307, doi:10.3390/su8040307. <http://www.mdpi.com/2071-1050/8/4/307>
24 (Accessed December 30, 2017).
- 25 Nyong, A., F. Adesina, and B. Osman Elasha, 2007a: The value of indigenous knowledge in climate
26 change mitigation and adaptation strategies in the African Sahel. *Mitig. Adapt. Strateg. Glob.*
27 *Chang.*, **12**, 787–797, doi:10.1007/s11027-007-9099-0.
28 <http://link.springer.com/10.1007/s11027-007-9099-0> (Accessed May 18, 2018).
- 29 —, —, and —, 2007b: The value of indigenous knowledge in climate change mitigation and
30 adaptation strategies in the African Sahel. *Mitig. Adapt. Strateg. Glob. Chang.*, **12**, 787–797,
31 doi:10.1007/s11027-007-9099-0. <http://link.springer.com/10.1007/s11027-007-9099-0>
32 (Accessed December 30, 2017).
- 33 Nyström, M., and Coauthors, 2012: Confronting Feedbacks of Degraded Marine Ecosystems.
34 *Ecosystems*, **15**, 695–710, doi:10.1007/s10021-012-9530-6.
35 <http://link.springer.com/10.1007/s10021-012-9530-6> (Accessed May 25, 2018).
- 36 O’ Brien, K., and Coauthors, 2012: Toward a sustainable and resilient future. *Managing the Risks of*
37 *Extreme Events and Disasters to Advance Climate Change Adaptation: Special Report of the*
38 *Intergovernmental Panel on Climate Change*, Vol. 9781107025 of, 437–486.
- 39 O’Brien, C.O., Scott, Z., Smith, G., Barca, V., Kardan, A., Holmes, R. Watson, C., 2018: *Shock-*
40 *Responsive Social Protection Systems Research Synthesis REport*.
- 41 O’Hare, P., I. White, and A. Connelly, 2016: Insurance as maladaptation: Resilience and the ‘business
42 as usual’ paradox. *Environ. Plan. C Gov. Policy*, **34**, 1175–1193,
43 doi:10.1177/0263774X15602022. <http://journals.sagepub.com/doi/10.1177/0263774X15602022>
44 (Accessed February 26, 2019).
- 45 O’Neill, B. C., and Coauthors, 2017a: IPCC reasons for concern regarding climate change risks. *Nat.*
46 *Clim. Chang.*, doi:10.1038/nclimate3179.
- 47 —, and Coauthors, 2017b: IPCC reasons for concern regarding climate change risks. *Nat. Clim.*

- 1 *Chang.*, 7, 28–37, doi:10.1038/nclimate3179.
2 <http://www.nature.com/doi/10.1038/nclimate3179> (Accessed December 15, 2017).
- 3 Oakes, L. E., N. M. Ardoin, and E. F. Lambin, 2016: “I know, therefore I
4 adapt?” Complexities of individual adaptation to climate-induced forest dieback in
5 Alaska. *Ecol. Soc.*, 21, art40, doi:10.5751/ES-08464-210240.
6 <http://www.ecologyandsociety.org/vol21/iss2/art40/> (Accessed April 14, 2019).
- 7 Oba, G., 1994: THE IMPORTANCE OF PASTORALISTS' INDIGENOUS COPING STRATEGIES
8 FOR PLANNING DROUGHT MANAGEMENT IN THE ARID ZONE OF KENYA. *Nomad.*
9 *People.*, 5, 89–119, doi:10.2307/43123620. <http://www.jstor.org/stable/43123620> (Accessed
10 May 22, 2018).
- 11 —, 2013: The sustainability of pastoral production in Africa. 54–61, doi:10.4324/9780203105979-
12 11. [https://www.taylorfrancis.com/books/e/9781136255854/chapters/10.4324/9780203105979-
13 11](https://www.taylorfrancis.com/books/e/9781136255854/chapters/10.4324/9780203105979-11) (Accessed April 7, 2019).
- 14 Oberlack, C., 2017: Diagnosing institutional barriers and opportunities for adaptation to climate
15 change. *Mitig. Adapt. Strateg. Glob. Chang.*, doi:10.1007/s11027-015-9699-z.
- 16 Ochieng, C., S. Juhola, and F. X. Johnson, 2014: The societal role of charcoal production in climate
17 change adaptation of the arid and semi-arid lands (ASALs) of Kenya. *Climate change*
18 *adaptation and development: transforming paradigms and practices*, T.H. Inderberg, S.H.
19 Eriksen, K.L. O'Brien, and L. Sygna, Eds., Routledge, London [http://su.diva-
20 portal.org.ezp.sub.su.se/smash/record.jsf?pid=diva2%3A788097&dswid=-3459](http://su.diva-portal.org.ezp.sub.su.se/smash/record.jsf?pid=diva2%3A788097&dswid=-3459) (Accessed April
21 14, 2019).
- 22 Ockwell, D., A. Sagar, and H. de Coninck, 2015: Collaborative research and development
23 (R&D) for climate technology transfer and uptake in developing countries: towards a needs
24 driven approach. *Clim. Change*, 131, 401–415, doi:10.1007/s10584-014-1123-2.
25 <http://link.springer.com/10.1007/s10584-014-1123-2> (Accessed May 22, 2018).
- 26 OECD, 2014: *Social Institutions and Gender Index*. Paris, France,.
- 27 —, 2015: Climate Finance in 2013-14 and the USD 100 billion goal. *World Econ. Forum*,
28 doi:<http://dx.doi.org/10.1787/9789264249424-en>.
- 29 —, 2018: *Joint Working Party on Agriculture and the Environment: A Global Economic*
30 *Evaluation Of GHG Mitigation Policies For Agriculture*. Paris,.
- 31 Oels, A., 2013: Rendering climate change governable by risk: From probability to contingency.
32 *Geoforum*, 45, 17–29, doi:10.1016/j.geoforum.2011.09.007.
- 33 Oglethorpe, J., J. Ericson, R. . Bilsborrow, and J. Edmond, 2007: People on the Move: Reducing the
34 Impact of Human Migration on Biodiversity. *World Wildl. Fund Conserv. Int. Found.*,
35 doi:10.13140/2.1.2987.0083.
36 [https://www.researchgate.net/publication/267327782_People_on_the_Move_Reducing_the_Imp
37 act_of_Human_Migration_on_Biodiversity](https://www.researchgate.net/publication/267327782_People_on_the_Move_Reducing_the_Impact_of_Human_Migration_on_Biodiversity) (Accessed May 22, 2018).
- 38 Ojea, E., 2015: Challenges for mainstreaming Ecosystem-based Adaptation into the international
39 climate agenda. *Curr. Opin. Environ. Sustain.*, 14, 41–48, doi:10.1016/j.cosust.2015.03.006.
- 40 Oke, T. R., G. Mills, A. Christen, and J. A. Voogt, 2017: *Urban climates*.
- 41 Oladele, S. O., A. J. Adeyemo, and M. A. Awodun, 2019: Influence of rice husk biochar and
42 inorganic fertilizer on soil nutrients availability and rain-fed rice yield in two contrasting soils.
43 *Geoderma*, 336, 1–11, doi:10.1016/J.GEODERMA.2018.08.025. [https://www.sciencedirect-
44 com.ezp.sub.su.se/science/article/pii/S0016706118303653?via%3Dihub](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0016706118303653?via%3Dihub) (Accessed April 15,
45 2019).
- 46 Oliveira Júnior, J. G. C., R. J. Ladle, R. Correia, and V. S. Batista, 2016: Measuring what matters –
47 Identifying indicators of success for Brazilian marine protected areas. *Mar. Policy*,
48 doi:10.1016/j.marpol.2016.09.018.

- 1 Oliver, D. M., R. D. Fish, M. Winter, C. J. Hodgson, A. L. Heathwaite, and D. R. Chadwick, 2012:
2 Valuing local knowledge as a source of expert data: Farmer engagement and the design of
3 decision support systems. *Environ. Model. Softw.*, **36**, 76–85,
4 doi:10.1016/J.ENVSOF.2011.09.013.
5 <https://www.sciencedirect.com/science/article/pii/S1364815211002118> (Accessed April 9,
6 2019).
- 7 Olsson, A., S. Grö Nkvist, M. Rten Lind, and J. Yan, 2016: The elephant in the room – A comparative
8 study of uncertainties in carbon offsets. doi:10.1016/j.envsci.2015.11.004. [https://ac-els-cdn-
9 com.ezp.sub.su.se/S1462901115301052/1-s2.0-S1462901115301052-main.pdf?_tid=53efa47e-
10 5a81-44b3-b913-525206d5918f&acdnat=1527022084_bf39aa58dded8f9904de9ace7eb0c428](https://ac-els-cdn-com.ezp.sub.su.se/S1462901115301052/1-s2.0-S1462901115301052-main.pdf?_tid=53efa47e-5a81-44b3-b913-525206d5918f&acdnat=1527022084_bf39aa58dded8f9904de9ace7eb0c428)
11 (Accessed May 22, 2018).
- 12 Olsson, L., and Coauthors, 2014: Livelihoods and Poverty. *Climate Change 2014: Impacts,*
13 *Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working*
14 *Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change,*
15 T.E.B. Field, C.B., V.R. Barros, D.J. Dokken, K.J. Mach, M.D. Mastrandrea, S.M. M.
16 Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova, B. Girma, E.S. Kissel, A.N. Levy, and And
17 L.L.W. P.R. Mastrandrea, Eds., Cambridge University Press, Cambridge and New York
18 https://www.ipcc.ch/site/assets/uploads/2018/02/WGIIAR5-Chap13_FINAL.pdf (Accessed
19 April 9, 2019).
- 20 —, and Coauthors, 2015a: Livelihoods and poverty. *Climate Change 2014 Impacts, Adaptation and*
21 *Vulnerability: Part A: Global and Sectoral Aspects*, 793–832.
- 22 —, A. Jerneck, H. Thoren, J. Persson, and D. O’Byrne, 2015b: Why resilience is unappealing to
23 social science: Theoretical and empirical investigations of the scientific use of resilience. *Sci.*
24 *Adv.*, doi:10.1126/sciadv.1400217.
- 25 Olsson, P., L. H. Gunderson, S. R. Carpenter, P. Ryan, L. Lebel, C. Folke, and C. S. Holling, 2006:
26 Shooting the rapids: Navigating transitions to adaptive governance of social-ecological systems.
27 *Ecol. Soc.*, **11**, doi:18.
- 28 Onibon, A., B. Dabiré, and L. Ferroukhi, 1999: Local practices and the decentralization and
29 devolution of natural resource management in French-speaking West Africa. *Unasylva*,.
- 30 Oppenheimer, M., and Coauthors, 2014: Emergent risks and key vulnerabilities. *Climate Change*
31 *2014 Impacts, Adaptation and Vulnerability: Part A: Global and Sectoral Aspects*, 1039–1100.
- 32 Organization for Economic Cooperation and Development, 2008: *Economic aspects of adaptation to*
33 *climate change: Costs, benefits and policy instruments*. 133 pp.
- 34 Orlove, B., C. Roncoli, M. Kabugo, and A. Majugu, 2010: Indigenous climate knowledge in southern
35 Uganda: the multiple components of a dynamic regional system. *Clim. Change*, **100**, 243–265,
36 doi:10.1007/s10584-009-9586-2. <http://link.springer.com/10.1007/s10584-009-9586-2>
37 (Accessed April 7, 2019).
- 38 Orlowsky, B., and S. I. Seneviratne, 2012: Global changes in extreme events: regional and seasonal
39 dimension. *Clim. Change*, **110**, 669–696, doi:10.1007/s10584-011-0122-9.
40 <http://link.springer.com/10.1007/s10584-011-0122-9> (Accessed April 9, 2019).
- 41 Orr, A. L., and Coauthors, 2017: *Scientific Conceptual Framework for Land Degradation Neutrality.*
42 *A Report of the Science-Policy Interface.* Bonn, Germany,
43 http://catalogue.unccd.int/814_LDN_CF_report_web-english.pdf.
- 44 Osei-Tutu, P., M. Pregernig, and B. Pokorny, 2014: Legitimacy of informal institutions in
45 contemporary local forest management: insights from Ghana. *Biodivers. Conserv.*, **23**, 3587–
46 3605, doi:10.1007/s10531-014-0801-8.
- 47 Ostrom, E., 2010: Beyond markets and states: Polycentric governance of complex economic systems.
48 *Am. Econ. Rev.*, **100**, 641–672, doi:10.1257/aer.100.3.641.

- 1 —, 2011: Background on the Institutional Analysis and. *Policy Stud. J.*, **39**, 7–27,
2 doi:10.1111/j.1541-0072.2010.00394.x.
- 3 —, 2012: Nested externalities and polycentric institutions: Must we wait for global solutions to
4 climate change before taking actions at other scales? *Econ. Theory*, doi:10.1007/s00199-010-
5 0558-6.
- 6 Oteros-Rozas, E., R. Ontillera-Sánchez, P. Sanosa, E. Gómez-Baggethun, V. Reyes-García, and J. A.
7 González, 2013: Traditional ecological knowledge among transhumant pastoralists in
8 Mediterranean Spain. *Ecol. Soc.*, **18**, art33, doi:10.5751/ES-05597-180333.
9 <http://www.ecologyandsociety.org/vol18/iss3/art33/> (Accessed April 7, 2019).
- 10 Otto, F. E. L., and Coauthors, 2015: Explaining extreme events of 2014 from a climate perspective:
11 Factors other than climate change, main drivers of 2014/2015 water shortage in Southeast
12 Brazil. *Bull. Am. Meteorol. Soc.*, **96**, S35–S40, doi:10.1175/BAMS-D-15-00120.1.
- 13 Van Oudenhoven, A. P. E., K. Petz, R. Alkemade, L. Hein, and R. S. de Groot, 2012: Framework for
14 systematic indicator selection to assess effects of land management on ecosystem services. *Ecol.*
15 *Indic.*, **21**, 110–122.
- 16 Outka, U., 2012: Environmental Law and Fossil Fuels: Barriers to Renewable Energy. *Vanderbilt Law*
17 *Rev.*, **65**, 1679–1721.
- 18 Outreville, J. F., 2011a: The relationship between insurance growth and economic development - 80
19 empirical papers for a review of the literature. *ICER Work. Pap.*,
20 <https://ideas.repec.org/p/icr/wpicer/12-2011.html> (Accessed April 3, 2019).
- 21 —, 2011b: The relationship between insurance growth and economic development - 80 empirical
22 papers for a review of the literature. *ICER Work. Pap.*,
- 23 Owen, R., G. Brennan, and F. Lyon, 2018: Enabling investment for the transition to a low carbon
24 economy: government policy to finance early stage green innovation. *Curr. Opin. Environ.*
25 *Sustain.*, **31**, 137–145, doi:10.1016/j.cosust.2018.03.004.
26 <https://doi.org/10.1016/j.cosust.2018.03.004>.
- 27 Pacetti, T., E. Caporali, and M. C. Rulli, 2017: Floods and food security: A method to estimate the
28 effect of inundation on crops availability. *Adv. Water Resour.*,
29 doi:10.1016/j.advwatres.2017.06.019.
- 30 Pachauri, S., and L. Jiang, 2008: The household energy transition in India and China. *Energy Policy*,
31 **36**, 4022–4035, doi:10.1016/j.enpol.2008.06.016. [https://www.sciencedirect-](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0301421508003029)
32 [com.ezp.sub.su.se/science/article/pii/S0301421508003029](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0301421508003029) (Accessed April 14, 2019).
- 33 —, N. D. Rao, and C. Cameron, 2018: Outlook for modern cooking energy access in Central
34 America. *PLoS One*, **13**, e0197974, doi:10.1371/journal.pone.0197974.
35 <https://dx.plos.org/10.1371/journal.pone.0197974> (Accessed November 16, 2018).
- 36 Pacheco, P., D. Barry, P. Cronkleton, and A. Larson, 2012: The Recognition of Forest Rights in Latin
37 America: Progress and Shortcomings of Forest Tenure Reforms. *Soc. Nat. Resour.*, **25**, 556–571,
38 doi:10.1080/08941920.2011.574314.
- 39 —, R. Pocard-Chapuis, I. Garcia Drigo, M.-G. Piketty, and M. Thales, 2016: Linking sustainable
40 production and enhanced landscape governance in the Amazon: towards territorial certification
41 (TerraCert). <http://agritrop.cirad.fr/583047/> (Accessed May 23, 2018).
- 42 Pagdee, A., Y. S. Kim, and P. J. Daugherty, 2006: What makes community forest management
43 successful: A meta-study from community forests throughout the world. *Soc. Nat. Resour.*, **19**,
44 33–52, doi:10.1080/08941920500323260.
- 45 Pahl-Wostl, C., 2009: A conceptual framework for analysing adaptive capacity and multi-level
46 learning processes in resource governance regimes. *Glob. Environ. Chang.*, **19**, 354–365,
47 doi:10.1016/j.gloenvcha.2009.06.001.

- 1 —, 2017a: Governance of the water-energy-food security nexus: A multi-level coordination
2 challenge. *Environmental Science and Policy*.
- 3 —, 2017b: An Evolutionary Perspective on Water Governance: From Understanding to
4 Transformation. *Water Resour. Manag.*, **31**, 2917–2932, doi:10.1007/s11269-017-1727-1.
- 5 —, J. Sendzimir, P. Jeffrey, J. Aerts, G. Berkamp, and K. Cross, 2007: Managing change toward
6 adaptive water management through social learning. *Ecol. Soc.*, **12**, doi:30.
- 7 —, C. Vörösmarty, A. Bhaduri, J. Bogardi, J. Rockström, and J. Alcamo, 2013: Towards a
8 sustainable water future: shaping the next decade of global water research. *Curr. Opin. Environ.
9 Sustain.*, **5**, 708–714, doi:https://doi.org/10.1016/j.cosust.2013.10.012.
10 <http://www.sciencedirect.com/science/article/pii/S1877343513001425>.
- 11 —, A. Bhaduri, and A. Bruns, 2018a: Editorial special issue: The Nexus of water, energy and food
12 – An environmental governance perspective. *Environ. Sci. Policy*,
13 doi:10.1016/j.envsci.2018.06.021.
- 14 —, —, and —, 2018b: Editorial special issue: The Nexus of water, energy and food – An
15 environmental governance perspective. *Environmental Science and Policy*.
- 16 Palmer, J. R., 2014: Biofuels and the politics of land-use change: Tracing the interactions of discourse
17 and place in European policy making. *Environ. Plan. A*, doi:10.1068/a4684.
- 18 Palomo, I., M. R. Felipe-Lucia, E. M. Bennett, B. Martín-López, and U. Pascual, 2016: *Disentangling
19 the Pathways and Effects of Ecosystem Service Co-Production*. 1st ed. Elsevier Ltd., 245-283
20 pp. <http://dx.doi.org/10.1016/bs.aecr.2015.09.003>.
- 21 Pandey, A., and Coauthors, 2017: Aerosol emissions factors from traditional biomass cookstoves in
22 India: insights from field measurements. *Atmos. Chem. Phys.*, **17**, 13721–13729,
23 doi:10.5194/acp-17-13721-2017. <https://www.atmos-chem-phys.net/17/13721/2017/> (Accessed
24 April 14, 2019).
- 25 Pannell, D., 2008: Public Benefits, Private Benefits, and Policy Mechanism Choice for Land-Use
26 Change for Environmental Benefits. *Land Econ.*, **84**, 225–240.
- 27 Panteli, M., and P. Mancarella, 2015: Influence of extreme weather and climate change on the
28 resilience of power systems: Impacts and possible mitigation strategies. *Electr. Power Syst. Res.*,
29 **127**, 259–270, doi:10.1016/j.epsr.2015.06.012.
- 30 Papaioannou, K. J., 2016: Climate shocks and conflict: Evidence from colonial Nigeria. *Polit. Geogr.*,
31 **50**, 33–47, doi:10.1016/j.polgeo.2015.07.001.
- 32 Papatoma-Köhle, M., C. Promper, and T. Glade, 2016: A Common Methodology for Risk
33 Assessment and Mapping of Climate Change Related Hazards—Implications for Climate
34 Change Adaptation Policies. *Climate*, doi:10.3390/cli4010008.
- 35 Park, S. E., N. . Marshall, E. Jakku, A. . Dowd, S. . Howden, E. Mendham, and A. Fleming, 2012:
36 Informing adaptation responses through theories of transformation. *Glob. Environ. Chang.*, **22**,
37 115–126.
- 38 Parnell, S., and R. Walawege, 2011: Sub-Saharan African urbanisation and global environmental
39 change. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2011.09.014.
- 40 Pascual, U., and Coauthors, 2017: Valuing nature’s contributions to people: the IPBES approach.
41 *Curr. Opin. Environ. Sustain.*, **26–27**, 7–16, doi:10.1016/j.cosust.2016.12.006.
- 42 Pathirana, A., Radhakrishnan, M., Ashley, R. et al, 2018: Climatic Change (2018) 149: 57.
43 <https://doi.org/10.1007/s10584-017-2059-0>. *Clim. Change*, **149**, 57–74.
44 <https://doi.org/10.1007/s10584-017-2059-0>.
- 45 Pathirana, A., Radhakrishnan, M., Quan, N. H. et al., 2018: Managing urban water systems with
46 significant adaptation deficits – unified framework for secondary cities : Part II - conceptual
47 framework. *Clim. Change*, **149**, 43–56.

- 1 Patterson, J., and Coauthors, 2017: Exploring the governance and politics of transformations towards
2 sustainability. *Environ. Innov. Soc. Transitions*, **24**, 1–16, doi:10.1016/j.eist.2016.09.001.
3 <http://linkinghub.elsevier.com/retrieve/pii/S2210422416300843> (Accessed May 26, 2018).
- 4 Patz, J. A., and Coauthors, 2004: Unhealthy landscapes: Policy recommendations on land use change
5 and infectious disease emergence. *Environ. Health Perspect.*, **112**, 1092–1098,
6 doi:10.1289/EHP.6877. <http://www.ncbi.nlm.nih.gov/pubmed/15238283> (Accessed May 22,
7 2018).
- 8 Paul, S., S. Ghosh, K. Rajendran, and R. Murtugudde, 2018: Moisture Supply From the Western
9 Ghats Forests to Water Deficit East Coast of India. *Geophys. Res. Lett.*, **45**, 4337–4344.
- 10 Paveglio, T. B., C. Kooistra, T. Hall, and M. Pickering, 2016: Residents ' Well-Being : What Factors
11 Influence. **62**, 59–69.
- 12 Payne, G., 2001: Urban land tenure policy options: titles or rights? *Habitat Int.*, **3**, 415–429.
13 [https://www.infona.pl/resource/bwmeta1.element.elsevier-f73bf0a7-8135-368b-bd9d-
14 1f65e75b167b](https://www.infona.pl/resource/bwmeta1.element.elsevier-f73bf0a7-8135-368b-bd9d-1f65e75b167b) (Accessed November 3, 2018).
- 15 Pearson, T. R. H., S. Brown, L. Murray, and G. Sidman, 2017: Greenhouse gas emissions from
16 tropical forest degradation: an underestimated source. *Carbon Balance Manag.*, **12**, 3,
17 doi:10.1186/s13021-017-0072-2.
18 <https://cbmjournals.biomedcentral.com/articles/10.1186/s13021-017-0072-2> (Accessed April 15,
19 2019).
- 20 Pecl, G. T., and Coauthors, 2017: Biodiversity redistribution under climate change: Impacts on
21 ecosystems and human well-being. *Science (80-.)*, doi:10.1126/science.aai9214.
- 22 Peel, J., and H. M. Osofsky, 2017: A Rights Turn in Climate Change Litigation?†. *Transnational
23 Environmental Law*.
- 24 Pelling, M., 2010: *Adaptation to climate change: From resilience to transformation*.
25 ———, 2011: *Adaptation to climate change : from resilience to transformation*. Routledge, 203 pp.
26 [https://www.routledge.com/Adaptation-to-Climate-Change-From-Resilience-to-Transformation-
27 1st-Edition/Pelling/p/book/9780415477505](https://www.routledge.com/Adaptation-to-Climate-Change-From-Resilience-to-Transformation-1st-Edition/Pelling/p/book/9780415477505) (Accessed April 3, 2019).
- 28 ———, and B. Wisner, 2012: African cities of hope and risk. *Disaster Risk Reduction: Cases from
29 Urban Africa*.
- 30 Pendrill, F., M. Persson, J. Godar, and T. Kastner, 2019: Deforestation displaced: trade in forest-risk
31 commodities and the prospects for a global forest transition. *Environ. Res. Lett.*,
32 doi:10.1088/1748-9326/ab0d41. <http://iopscience.iop.org/article/10.1088/1748-9326/ab0d41>
33 (Accessed April 8, 2019).
- 34 Pereira, H. M., and Coauthors, 2010: Scenarios for global biodiversity in the 21st century. *Science
35 (80-.)*, **330**, 1496–1501, doi:10.1126/science.1196624.
- 36 Perera, N., E. Boyd Gill Wilkins, and R. Phillips Itty, 2015: *Literature Review on Energy Access and
37 Adaptation to Climate Change*. London, 89 pp.
38 [https://assets.publishing.service.gov.uk/media/57a0896b40f0b652dd0001fe/LitRev-
39 EnergyAccessandAdaptation-Final-2.pdf](https://assets.publishing.service.gov.uk/media/57a0896b40f0b652dd0001fe/LitRev-EnergyAccessandAdaptation-Final-2.pdf) (Accessed April 14, 2019).
- 40 Pérez, I., M. A. Janssen, and J. M. Anderies, 2016: Food security in the face of climate change:
41 Adaptive capacity of small-scale social-ecological systems to environmental variability. *Glob.
42 Environ. Chang.*, **40**, 82–91, doi:https://doi.org/10.1016/j.gloenvcha.2016.07.005.
- 43 Perry, J., 2015: Climate change adaptation in the world's best places: A wicked problem in need of
44 immediate attention. *Landsc. Urban Plan.*, **133**, 1–11.
- 45 Persha, L., and K. Andersson, 2014: Elite capture risk and mitigation in decentralized forest
46 governance regimes. *Glob. Environ. Chang.*, **24**, 265–276,
47 doi:10.1016/J.GLOENVCHA.2013.12.005.

- 1 <https://www.sciencedirect.com/science/article/pii/S0959378013002355> (Accessed October 3,
2 2018).
- 3 Persha, L., A. Agrawal, and A. Chhatre, 2011: Social and Ecological Synergy: Local Rulemaking,
4 Forest Livelihoods, and Biodiversity Conservation. *Science* (80-.), **331**, 1606–1608,
5 doi:10.1126/science.1199343. <http://www.sciencemag.org/cgi/doi/10.1126/science.1199343>
6 (Accessed April 9, 2019).
- 7 Persson, J., E. L. Johansson, and L. Olsson, 2018: Harnessing local knowledge for scientific
8 knowledge production: challenges and pitfalls within evidence-based sustainability studies.
9 *Ecol. Soc.*, **23**, art38, doi:10.5751/ES-10608-230438.
10 <https://www.ecologyandsociety.org/vol23/iss4/art38/> (Accessed April 8, 2019).
- 11 Pert, P. L., R. Hill, K. Maclean, A. Dale, P. Rist, J. Schmider, L. Talbot, and L. Tawake, 2015:
12 Mapping cultural ecosystem services with rainforest aboriginal peoples: Integrating biocultural
13 diversity, governance and social variation. *Ecosyst. Serv.*, doi:10.1016/j.ecoser.2014.10.012.
- 14 Von Peter, G., S. Von Dahlen, and S. Saxena, 2012: *Unmitigated disasters? New evidence on the*
15 *macroeconomic cost of natural catastrophes*. www.bis.org (Accessed April 3, 2019).
- 16 Pezzey, J. C. V., 2019: Why the social cost of carbon will always be disputed. *Wiley Interdiscip. Rev.*
17 *Clim. Chang.*, **10**, 1–12, doi:10.1002/wcc.558.
- 18 Pielke, R. A., G. Marland, R. A. Betts, T. N. Chase, J. L. Eastman, J. O. Niles, D. d. S. Niyogi, and S.
19 W. Running, 2002: The influence of land-use change and landscape dynamics on the climate
20 system: relevance to climate-change policy beyond the radiative effect of greenhouse gases.
21 *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, **360**, 1705–1719, doi:10.1098/rsta.2002.1027.
22 <http://www.ncbi.nlm.nih.gov/pubmed/12460493> (Accessed May 18, 2018).
- 23 Pieraccini, M., 2015: Rethinking participation in environmental decision-making: Epistemologies of
24 marine conservation in South-East England. *J. Environ. Law*, **27**, 45–67, doi:10.1093/jel/equ035.
- 25 Pierson, F. B., and C. J. Williams, 2016: Ecohydrologic Impacts of Rangeland Fire on Runoff and
26 Erosion: A Literature Synthesis. *US Dep. Agric. For. Serv. Rocky Mt. Res. Station. Gen. Tech.*
27 *Rep. RMRS-GTR-351.*, 110 p.
- 28 Pierson, F. B., C. J. Williams, S. P. Hardegree, M. A. Wertz, J. J. Stone, and P. E. Clark, 2011: Fire,
29 plant invasions, and erosion events on Western Rangelands. *Rangel. Ecol. Manag.*,
30 doi:10.2111/REM-D-09-00147.1.
- 31 Pingali, P., 2015: Agricultural policy and nutrition outcomes – getting beyond the preoccupation with
32 staple grains. *Food Secur.*, doi:10.1007/s12571-015-0461-x.
- 33 Pingali, P. L., 2012: Green Revolution: Impacts, limits, and the path ahead. *Proc. Natl. Acad. Sci.*,
34 doi:10.1073/pnas.0912953109.
- 35 Pinho, P., R. Maia, and A. Monterroso, 2007: The quality of Portuguese Environmental Impact
36 Studies: The case of small hydropower projects. *Environ. Impact Assess. Rev.*, **27**, 189–205.
- 37 Pistorius, T., S. Reinecke, and A. Carrapatoso, 2017: A historical institutionalist view on merging
38 LULUCF and REDD+ in a post-2020 climate agreement. *Int. Environ. Agreements Polit. Law*
39 *Econ.*, **17**, 623–638, doi:10.1007/s10784-016-9330-0. [http://link.springer.com/10.1007/s10784-](http://link.springer.com/10.1007/s10784-016-9330-0)
40 [016-9330-0](http://link.springer.com/10.1007/s10784-016-9330-0) (Accessed May 22, 2018).
- 41 Pittelkow, C. M., and Coauthors, 2015: Productivity limits and potentials of the principles of
42 conservation agriculture. *Nature*, **517**, doi:10.1038/nature13809.
- 43 Plambeck, E. L., C. Hope, and J. Anderson, 1997: The Page95 model: Integrating the science and
44 economics of global warming. *Energy Econ.*, **19**, 77–101, doi:http://dx.doi.org/10.1016/S0140-
45 9883(96)01008-0.
- 46 Plummer, R., and J. Baird, 2013: Adaptive co-management for climate change adaptation:
47 Considerations for the barents region. *Sustain.*, **5**, 629–642, doi:10.3390/su5020629.

- 1 Poff, N. L., and Coauthors, 2003: River flows and water wars: emerging science for environmental
2 decision making. *Front. Ecol. Environ.*, **1**, 298–306, doi:10.1890/1540-
3 9295(2003)001[0298:RFAWWE]2.0.CO;2. [http://www.esajournals.org/doi/abs/10.1890/1540-](http://www.esajournals.org/doi/abs/10.1890/1540-9295(2003)001%5B0298:RFAWWE%5D2.0.CO;2)
4 9295(2003)001%5B0298:RFAWWE%5D2.0.CO;2 (Accessed May 24, 2018).
- 5 Pokharel, P., and S. X. Chang, 2019: Manure pellet, woodchip and their biochars differently affect
6 wheat yield and carbon dioxide emission from bulk and rhizosphere soils. *Sci. Total Environ.*,
7 **659**, 463–472, doi:10.1016/J.SCITOTENV.2018.12.380. [https://www.sciencedirect-](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0048969718352744?via%3Dihub)
8 com.ezp.sub.su.se/science/article/pii/S0048969718352744?via%3Dihub (Accessed April 15,
9 2019).
- 10 Popp, A., and Coauthors, 2014a: Land-use transition for bioenergy and climate stabilization: Model
11 comparison of drivers, impacts and interactions with other land use based mitigation options.
12 *Clim. Change*, doi:10.1007/s10584-013-0926-x.
- 13 ———, and Coauthors, 2017: Land-use futures in the shared socio-economic pathways. *Glob. Environ.*
14 *Chang.*, **42**, 331–345, doi:10.1016/J.GLOENVCHA.2016.10.002.
- 15 Popp, J., K. Peto, and J. Nagy, 2013: Pesticide productivity and food security. A review. *Agron.*
16 *Sustain. Dev.*, doi:10.1007/s13593-012-0105-x.
- 17 Popp, J., Z. Lakner, M. Harangi-Rákos, and M. Fári, 2014b: The effect of bioenergy expansion: Food,
18 energy, and environment. *Renew. Sustain. Energy Rev.*, doi:10.1016/j.rser.2014.01.056.
- 19 Porras, I., and N. Asquith, 2018: *Ecosystems, poverty alleviation and conditional transfers Guidance*
20 *for practitioners*. IIED, London, www.iied.org (Accessed November 1, 2018).
- 21 Porter, J. R., L. Xie, A. J. Challinor, K. Cochrane, S. M. Howden, M. M. Iqbal, D. B. Lobell, and M.
22 I. Travasso, 2014: Food security and food production systems. *Climate Change 2014: Impacts,*
23 *Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working*
24 *Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change,*
25 485–533.
- 26 Poudyal, M., B. S. Ramamonjisoa, N. Hockley, O. S. Rakotonarivo, J. M. Gibbons, R.
27 Mandimbiniaina, A. Rasoamanana, and J. P. G. Jones, 2016: Can REDD+ social safeguards
28 reach the “right” people? Lessons from Madagascar. *Glob. Environ. Chang.*, **37**, 31–42,
29 doi:10.1016/j.gloenvcha.2016.01.004.
- 30 Premalatha, M., T. Abbasi, and S. A. Abbasi, 2014a: Wind energy: Increasing deployment, rising
31 environmental concerns. *Renew. Sustain. Energy Rev.*, **31**, 270–288.
- 32 ———, ———, and ———, 2014b: A critical view on the eco-friendliness of small hydroelectric
33 installations. *Sci. Total Environ.*, **481**, 638–643.
- 34 Priefer, C., J. Jörissen, and O. Frör, 2017: Pathways to Shape the Bioeconomy. *Resources*, **6**, 10,
35 doi:10.3390/resources6010010. <http://www.mdpi.com/2079-9276/6/1/10> (Accessed May 23,
36 2018).
- 37 Prober, S. M., and Coauthors, 2017: Informing climate adaptation pathways in multi-use woodland
38 landscapes using the values-rules-knowledge framework. *Agric. Ecosyst. Environ.*, **241**, 39–53,
39 doi:10.1016/j.agee.2017.02.021. <http://dx.doi.org/10.1016/j.agee.2017.02.021>.
- 40 Prokopy, L. S., K. Floress, D. Klotthor-Weinkauff, and A. Baumgart-Getz, 2008: Determinants of
41 agricultural best management practice adoption: Evidence from the literature. *J. Soil Water*
42 *Conserv.*, **63**, 300–311, doi:10.2489/jswc.63.5.300.
- 43 Prokopy, L. S., J. G. Arbuckle, A. P. Barnes, V. R. Haden, A. Hogan, M. T. Niles, and J. Tyndall,
44 2015: Farmers and Climate Change: A Cross-National Comparison of Beliefs and Risk
45 Perceptions in High-Income Countries. *Environ. Manage.*, **56**, 492–504, doi:10.1007/s00267-
46 015-0504-2. <http://link.springer.com/10.1007/s00267-015-0504-2> (Accessed May 29, 2018).
- 47 Public Safety Canada, 2017: *2016-2017 Evaluation of the Disaster Financial Assistance*
48 *Arrangements*. Ottawa Canada, <https://www.publicsafety.gc.ca/cnt/rsrscs/pblctns/vltn-dsstr-fnncl->

- 1 sstnc-2016-17/index-en.aspx (Accessed October 3, 2018).
- 2 Puma, M. J., S. Bose, S. Y. Chon, and B. I. Cook, 2015: Assessing the evolving fragility of the global
3 food system. *Environ. Res. Lett.*, doi:10.1088/1748-9326/10/2/024007.
- 4 Pungetti, G., G. Oviedo, and D. Hooke, 2012: *Sacred species and sites : advances in biocultural*
5 *conservation*. Cambridge University Press, 472 pp.
6 [https://books.google.co.uk/books?hl=en&lr=&id=14MhAwAAQBAJ&oi=fnd&pg=PR11&dq=S](https://books.google.co.uk/books?hl=en&lr=&id=14MhAwAAQBAJ&oi=fnd&pg=PR11&dq=Sacred+species+and+sites:+Advances+in+biocultural+conservation&ots=eO4fsxYezZ&sig=Nrc6dmXvHO2wUhGEqda1Wsx04fw#v=onepage&q=Sacred+species+and+sites%3A+Advances+in+biocultural+conservation&f=false)
7 [acred+species+and+sites:+Advances+in+biocultural+conservation&ots=eO4fsxYezZ&sig=Nrc6](https://books.google.co.uk/books?hl=en&lr=&id=14MhAwAAQBAJ&oi=fnd&pg=PR11&dq=Sacred+species+and+sites:+Advances+in+biocultural+conservation&ots=eO4fsxYezZ&sig=Nrc6dmXvHO2wUhGEqda1Wsx04fw#v=onepage&q=Sacred+species+and+sites%3A+Advances+in+biocultural+conservation&f=false)
8 [dmXvHO2wUhGEqda1Wsx04fw#v=onepage&q=Sacred species and sites%3A Advances in](https://books.google.co.uk/books?hl=en&lr=&id=14MhAwAAQBAJ&oi=fnd&pg=PR11&dq=Sacred+species+and+sites:+Advances+in+biocultural+conservation&ots=eO4fsxYezZ&sig=Nrc6dmXvHO2wUhGEqda1Wsx04fw#v=onepage&q=Sacred+species+and+sites%3A+Advances+in+biocultural+conservation&f=false)
9 [biocultural conservation&f=false](https://books.google.co.uk/books?hl=en&lr=&id=14MhAwAAQBAJ&oi=fnd&pg=PR11&dq=Sacred+species+and+sites:+Advances+in+biocultural+conservation&ots=eO4fsxYezZ&sig=Nrc6dmXvHO2wUhGEqda1Wsx04fw#v=onepage&q=Sacred+species+and+sites%3A+Advances+in+biocultural+conservation&f=false) (Accessed April 7, 2019).
- 10 Purkus, Alexandra; Gawel, Erik; Thrän, D., 2012: *Bioenergy governance between market and*
11 *government failures: A new institutional economics perspective*. Leipzig, 27 pp.
12 <http://hdl.handle.net/10419/64556> (Accessed April 11, 2019).
- 13 Purkus, A., E. Gawel, and D. Thrän, 2018: Addressing uncertainty in decarbonisation policy mixes –
14 Lessons learned from German and European bioenergy policy. *Energy Res. Soc. Sci.*, **33**, 82–94,
15 doi:10.1016/j.erss.2017.09.020. <http://dx.doi.org/10.1016/j.erss.2017.09.020> (Accessed April 12,
16 2019).
- 17 Pynegar, E. L., J. P. G. Jones, J. M. Gibbons, and N. M. Asquith, 2018: The effectiveness of
18 Payments for Ecosystem Services at delivering improvements in water quality: lessons for
19 experiments at the landscape scale. *PeerJ*, doi:10.7717/peerj.5753.
- 20 Qasim, S., R. P. Shrestha, G. P. Shivakoti, and N. K. Tripathi, 2011: Socio-economic determinants of
21 land degradation in Pishin sub-basin, Pakistan. *Int. J. Sustain. Dev. World Ecol.*, **18**, 48–54,
22 doi:10.1080/13504509.2011.543844.
- 23 Qin, Z., J. B. Dunn, H. Kwon, S. Mueller, and M. M. Wander, 2016: Soil carbon sequestration and
24 land use change associated with biofuel production: Empirical evidence. *GCB Bioenergy*,
25 doi:10.1111/gcbb.12237.
- 26 Quan, J., and N. Dyer, 2008: Climate change and land tenure: The implications of climate change for
27 land tenure and land policy. *FAO L. Tenure Work. Pap.*,
28 <ftp://ftp.fao.org/docrep/fao/011/aj332e/aj332e00.pdf%5Cnhttp://files/4335/2146.html> (Accessed
29 May 23, 2018).
- 30 —, L. O. Naess, A. Newsham, A. Siteo, and M. C. Fernandez, 2017: The political economy of
31 REDD+ in mozambique: Implications for climate compatible development. *Mak. Clim. Compat.*
32 *Dev. Happen*, 151–181, doi:10.4324/9781315621579. <https://eprints.soas.ac.uk/22587/>
33 (Accessed May 23, 2018).
- 34 Quatrini, S., and N. D. Crossman, 2018: Most finance to halt desertification also benefits multiple
35 ecosystem services: A key to unlock investments in Land Degradation Neutrality? *Ecosyst.*
36 *Serv.*, doi:10.1016/j.ecoser.2018.04.003.
- 37 Quay, R., 2010: Anticipatory Governance. *J. Am. Plan. Assoc.*, **76**, 496–511,
38 doi:10.1080/01944363.2010.508428.
- 39 Le Quesne, F., 2017: *The role of insurance in integrated disaster and climate risk management: Evidence and lessons learned*.
- 41 Quinn, A. K., and Coauthors, 2018a: An analysis of efforts to scale up clean household energy for
42 cooking around the world. *Energy Sustain. Dev.*, **46**, 1–10, doi:10.1016/j.esd.2018.06.011.
43 <http://creativecommons.org/licenses/by/4.0/> (Accessed April 15, 2019).
- 44 —, and Coauthors, 2018b: An analysis of efforts to scale up clean household energy for cooking
45 around the world. *Energy Sustain. Dev.*, **46**, 1–10, doi:10.1016/j.esd.2018.06.011.
- 46 Quirion, P., 2009: Historic versus output-based allocation of GHG tradable allowances: A
47 comparison. *Clim. Policy*, **9**, 575–592, doi:10.3763/cpol.2008.0618.

- 1 Quisumbing, A. R., R. Meinzen-Dick, T. L. Raney, A. Croppenstedt, J. A. Behrman, and A.
2 Peterman, 2014: Closing the Knowledge Gap on Gender in Agriculture. *Gender in Agriculture*,
3 Springer Netherlands, Dordrecht, 3–27 http://link.springer.com/10.1007/978-94-017-8616-4_1
4 (Accessed April 2, 2019).
- 5 Radeloff, V. C., and Coauthors, 2018: Rapid growth of the US wildland-urban interface raises
6 wildfire risk. *Proc. Natl. Acad. Sci.*, **115**, 3314–3319.
- 7 Radhakrishnan, M., Nguyen, H., Gersonius, B. et al., 2018: The planning and phasing of adaptation
8 responses are essential to tackle uncertainties and ensure positive outcomes while adapting to
9 changing circumstances. Understanding the evolution of coping and adaptation responses and
10 their capacities is a prerequi. *Clim. Change*, **149**, 29–41. [https://doi.org/10.1007/s10584-017-](https://doi.org/10.1007/s10584-017-1999-8)
11 [1999-8](https://doi.org/10.1007/s10584-017-1999-8).
- 12 Radhakrishnan, M., A. Pathirana, R. Ashley, and C. Zevenbergen, 2017: Structuring Climate
13 Adaptation through Multiple Perspectives: Framework and Case Study on Flood Risk
14 Management. *Water*, doi:10.3390/w9020129.
- 15 Rahman, M.M., Khan, M.N.I., Hoque, A.K.F., Ahmed, I., 2014: Carbon stockin the Sundarbans
16 mangrove forest: spatial variations in vegetation types and salinity zones. *Wetl. Ecol. Manag.*,.
- 17 Rahman, H. M. T., S. K. Sarker, G. M. Hickey, M. Mohasinul Haque, and N. Das, 2014: Informal
18 Institutional Responses to Government Interventions: Lessons from Madhupur National Park,
19 Bangladesh. *Environ. Manage.*, **54**, 1175–1189, doi:10.1007/s00267-014-0325-8.
- 20 Rahman, M. A., 2018: Governance matters: climate change, corruption, and livelihoods in
21 Bangladesh. *Clim. Change*, **147**, 313–326, doi:10.1007/s10584-018-2139-9.
22 <https://doi.org/10.1007/s10584-018-2139-9>.
- 23 Rajamani, L., 2011: The Cancun climate agreements: Reading the text, subtext and tea leaves. *Int.*
24 *Comp. Law Q.*, **60**, 499–519, doi:10.1017/S0020589311000078.
- 25 Raju, K. V, A. Aziz, S. S. M. Sundaram, M. Sekher, S. P. Wani, and T. K. Sreedevi, 2008: *Guidelines*
26 *for Planning and Implementation of Watershed Development Program in India: A*
27 *Review.Global Theme on Agroecosystems Report no. 48*. International Crops Research Institute
28 for the Semi-Arid Tropics, <http://oar.icrisat.org/2353/> (Accessed April 4, 2019).
- 29 Raleigh, C., H. J. Choi, and D. Kniveton, 2015: The devil is in the details: An investigation of the
30 relationships between conflict, food price and climate across Africa. *Glob. Environ. Chang.*,
31 doi:10.1016/j.gloenvcha.2015.03.005.
- 32 Ramnath, M., 2008: Surviving the Forest Rights Act: Between Scylla and Charybdis. *Econ. Polit.*
33 *Wkly.*, **43**, 37–42. [https://www.epw.in/journal/2008/09/perspectives/surviving-forest-rights-act-](https://www.epw.in/journal/2008/09/perspectives/surviving-forest-rights-act-between-scylla-and-charybdis.html)
34 [between-scylla-and-charybdis.html](https://www.epw.in/journal/2008/09/perspectives/surviving-forest-rights-act-between-scylla-and-charybdis.html) (Accessed November 3, 2018).
- 35 Ramos, J. M., 2014: Anticipatory governance: Traditions and trajectories for strategic design. *J.*
36 *Futur. Stud.*, **19**, 35–52.
- 37 Ranatunga, T., S. T. Y. Tong, Y. Sun, and Y. J. Yang, 2014: A total water management analysis of the
38 Las Vegas Wash watershed, Nevada. *Phys. Geogr.*, doi:10.1080/02723646.2014.908763.
- 39 Ranjay K. Singh, Shah M. Hussain, T. Riba, Anshuman Singh, Egul Padung, Orrik Rallen, Y. J. Lego,
40 Ajay Kumar Bhardwaj 2018. Classification and management of community forests in Indian
41 Eastern Himalayas: implications on ecosystem services, conservation and livelihoods.
42 Ecological Processes 7:27.Singh, S. P., and M. Swanson, 2017: How issue frames shape beliefs
43 about the importance of climate change policy across ideological and partisan groups. *PLoS*
44 *One*, **12**, doi:10.1371/journal.pone.0181401.
- 45 Rao, N., 2017a: Assets, Agency and Legitimacy: Towards a Relational Understanding of Gender
46 Equality Policy and Practice. *World Dev.*, **95**, 43–54, doi:10.1016/j.worlddev.2017.02.018.
47 <https://linkinghub.elsevier.com/retrieve/pii/S0305750X15308810> (Accessed April 2, 2019).
- 48 —, 2017b: Assets, Agency and Legitimacy: Towards a Relational Understanding of Gender

- 1 Equality Policy and Practice. *World Dev.*, **95**, 43–54, doi:10.1016/j.worlddev.2017.02.018.
- 2 Rao, N. D., and S. Pachauri, 2017: Energy access and living standards: some observations on recent
3 trends. *Environ. Res. Lett.*, **12**, 025011, doi:10.1088/1748-9326/aa5b0d.
4 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=2/a=025011?key=crossref.695ec63a546756c629d9924c69c1a863)
5 [9326/12/i=2/a=025011?key=crossref.695ec63a546756c629d9924c69c1a863](http://stacks.iop.org/1748-9326/12/i=2/a=025011?key=crossref.695ec63a546756c629d9924c69c1a863) (Accessed
6 November 16, 2018).
- 7 Rasul, G., 2014: Food, water, and energy security in South Asia: A nexus perspective from the Hindu
8 Kush Himalayan region{star, open}. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2014.01.010.
- 9 —, and B. Sharma, 2016: The nexus approach to water–energy–food security: an option for
10 adaptation to climate change. *Clim. Policy*, **16**, 682–702, doi:10.1080/14693062.2015.1029865.
11 <http://dx.doi.org/10.1080/14693062.2015.1029865>.
- 12 Rauken, T., P. K. Mydske, and M. Winsvold, 2014: Mainstreaming climate change adaptation at the
13 local level. *Local Environ.*, **20**, 408–423, doi:10.1080/13549839.2014.880412.
- 14 Ravera, F., I. Iniesta-Arandia, B. Martín-López, U. Pascual, and P. Bose, 2016: Gender perspectives
15 in resilience, vulnerability and adaptation to global environmental change. *Ambio*, **45**, 235–247,
16 doi:10.1007/s13280-016-0842-1. <http://link.springer.com/10.1007/s13280-016-0842-1>
17 (Accessed April 2, 2019).
- 18 Ray, B., and R. Shaw, 2016: Water Stress in the Megacity of Kolkata, India, and Its Implications for
19 Urban Resilience. *Urban Disasters and Resilience in Asia*, Elsevier Inc., 317–336.
- 20 Raza, W., and E. Poel, 2016: Impact and spill-over effects of an asset transfer program on
21 malnutrition: Evidence from a randomized control trial in Bangladesh.
22 doi:10.13140/RG.2.1.2283.0967.
- 23 Reddy, M. G., K. A. Kumar, P. T. Rao, and O. Springate-Baginski, 2011: Issues related to
24 Implementation of the Forest rights Act in Andhra Pradesh. *Econ. Polit. Wkly.*, 73–81.
- 25 Reed, M., and Coauthors, 2010: What is Social Learning? *Ecol. Soc.*, **15**, r1, doi:Article.
- 26 Reed, M. S., and Coauthors, 2014: Improving the link between payments and the provision of
27 ecosystem services in agri-environment schemes. *Ecosyst. Serv.*,
28 doi:10.1016/j.ecoser.2014.06.008.
- 29 Reichardt, K., K. S. Rogge, and S. Negro, 2015: Unpacking the policy processes for addressing
30 systemic problems: The case of the technological innovation system of offshore wind in
31 Germany. *Work. Pap. Sustain. Innov. No. S 02/2015*, doi:10.1017/CBO9781107415324.004.
- 32 Reid, H., 2016: Ecosystem- and community-based adaptation: learning from community-based
33 natural resource management management. *Clim. Dev.*, **8**, 4–9,
34 doi:10.1080/17565529.2015.1034233.
- 35 Reisinger, A., P. Havlik, K. Riahi, O. van Vliet, M. Obersteiner, and M. Herrero, 2013: Implications
36 of alternative metrics for global mitigation costs and greenhouse gas emissions from agriculture.
37 *Clim. Change*, **117**, 677–690, doi:10.1007/s10584-012-0593-3. [https://doi.org/10.1007/s10584-](https://doi.org/10.1007/s10584-012-0593-3)
38 [012-0593-3](https://doi.org/10.1007/s10584-012-0593-3).
- 39 Ren, Z., and Coauthors, 2016: Predicting malaria vector distribution under climate change scenarios
40 in China: Challenges for malaria elimination. *Sci. Rep.*, **6**, 20604, doi:10.1038/srep20604.
41 <http://www.nature.com/articles/srep20604> (Accessed December 29, 2017).
- 42 Renn, O., and P. Schweizer, 2009: Inclusive Risk Governance: Concepts and. *Environ. Policy Gov.*,
43 **19**, 174–185, doi:10.1002/eet.
- 44 Reyer, C. P. O., K. K. Rigaud, E. Fernandes, W. Hare, O. Serdeczny, and H. J. Schellnhuber, 2017:
45 Turn down the heat: regional climate change impacts on development. *Regional Environmental*
46 *Change*.
- 47 Riahi, K., and Coauthors, 2017: The Shared Socioeconomic Pathways and their energy, land use, and

- 1 greenhouse gas emissions implications: An overview. *Glob. Environ. Chang.*, **42**, 153–168,
2 doi:10.1016/J.GLOENVCHA.2016.05.009.
- 3 Richards, M., Bruun, T.B., Campbell, B.M., Gregersen, L.E., Huyer, S. et al., 2015: *How countries*
4 *plan to address agricultural adaptation and mitigation*. 1-8 pp.
- 5 Richards, M., and Coauthors, 2017: High-resolution spatial modelling of greenhouse gas emissions
6 from land-use change to energy crops in the United Kingdom. *GCB Bioenergy*,
7 doi:10.1111/gcbb.12360.
- 8 Richards, P., 1985: Indigenous agricultural revolution: ecology and food production in west Africa.
9 [http://www.sidalc.net/cgi-](http://www.sidalc.net/cgi-bin/wxis.exe/?IsisScript=UACHBC.xis&method=post&formato=2&cantidad=1&expresion=mf=053258)
10 [bin/wxis.exe/?IsisScript=UACHBC.xis&method=post&formato=2&cantidad=1&expresion=mf](http://www.sidalc.net/cgi-bin/wxis.exe/?IsisScript=UACHBC.xis&method=post&formato=2&cantidad=1&expresion=mf=053258)
11 [=053258](http://www.sidalc.net/cgi-bin/wxis.exe/?IsisScript=UACHBC.xis&method=post&formato=2&cantidad=1&expresion=mf=053258) (Accessed April 7, 2019).
- 12 Richter, B. D., and Coauthors, 2017: Opportunities for saving and reallocating agricultural water to
13 alleviate water scarcity. *Water Policy*, doi:10.2166/wp.2017.143.
- 14 Rietig, K., 2018: The Links Among Contested Knowledge, Beliefs, and Learning in European
15 Climate Governance: From Consensus to Conflict in Reforming Biofuels Policy. *Policy Stud. J.*,
16 **46**, 137–159, doi:10.1111/psj.12169. <http://doi.wiley.com/10.1111/psj.12169> (Accessed April
17 12, 2019).
- 18 Rights and Resources Initiative, 2015a: *Who owns the world's land? A global baseline of formally*
19 *recognized indigenous and community land rights*. Washington D C,
20 www.rightsandresources.org. (Accessed April 13, 2019).
- 21 —, 2015b: *Who owns the world's land? A global baseline of formally recognized indigenous and*
22 *community land rights*. Washington D C,.
- 23 —, 2018a: At a Crossroads: Consequential Trends in Recognition of Community-based Forest
24 Tenure From 2002-2017. *Rights Resour. Initiat.*,
25 [https://rightsandresources.org/en/publication/at-a-crossroads-trends-in-recognition-of-](https://rightsandresources.org/en/publication/at-a-crossroads-trends-in-recognition-of-community-based-forest-tenure-from-2002-2017/#.XLIQ3-gzY2w)
26 [community-based-forest-tenure-from-2002-2017/#.XLIQ3-gzY2w](https://rightsandresources.org/en/publication/at-a-crossroads-trends-in-recognition-of-community-based-forest-tenure-from-2002-2017/#.XLIQ3-gzY2w) (Accessed April 13, 2019).
- 27 —, 2018b: At a Crossroads: Consequential Trends in Recognition of Community-based Forest
28 Tenure From 2002-2017. *Rights Resour. Initiat.*,.
- 29 Rigon, A., 2014: Building Local Governance: Participation and Elite Capture in Slum-upgrading in
30 Kenya. *Dev. Change*, **45**, 257–283, doi:10.1111/dech.12078.
31 <http://doi.wiley.com/10.1111/dech.12078> (Accessed October 3, 2018).
- 32 Van Rijn, F., E. Bulte, and A. Adekunle, 2012: Social capital and agricultural innovation in Sub-
33 Saharan Africa. *Agric. Syst.*, **108**, 112–122, doi:10.1016/j.agsy.2011.12.003.
- 34 Riley, M., H. Sangster, H. Smith, R. Chiverrell, and J. Boyle, 2018: Will farmers work together for
35 conservation? The potential limits of farmers' cooperation in agri-environment measures. *Land*
36 *use policy*, **70**, 635–646, doi:10.1016/J.LANDUSEPOL.2017.10.049.
37 <https://www.sciencedirect.com/science/article/pii/S0264837717312759> (Accessed May 23,
38 2018).
- 39 Ring, I., and C. Schröter-Schlaack, 2011: *Instruments mixes for biodiversity policies*. 119-144 pp.
- 40 Ringler, E., A. Pašukonis, W. T. Fitch, L. Huber, W. Hödl, and M. Ringler, 2015: Flexible
41 compensation of uniparental care: female poison frogs take over when males disappear. *Behav.*
42 *Ecol.*, **26**, 1219–1225, doi:10.1093/beheco/arv069. [https://academic.oup.com/beheco/article-](https://academic.oup.com/beheco/article-lookup/doi/10.1093/beheco/arv069)
43 [lookup/doi/10.1093/beheco/arv069](https://academic.oup.com/beheco/article-lookup/doi/10.1093/beheco/arv069).
- 44 del Río, P., and E. Cerdá, 2017: The missing link: The influence of instruments and design features on
45 the interactions between climate and renewable electricity policies. *Energy Res. Soc. Sci.*, **33**,
46 49–58, doi:10.1016/j.erss.2017.09.010.
- 47 Rivera-Ferre, M. G., and Coauthors, 2016: Local agriculture traditional knowledge to ensure food

- 1 availability in a changing climate: revisiting water management practices in the Indo-Gangetic
2 Plains. *Agroecol. Sustain. Food Syst.*, doi:10.1080/21683565.2016.1215368.
- 3 Roberts, J. T., 2017: How will we pay for loss and damage? *Ethics, Policy Environ.*, **20**, 208–226.
- 4 Roberts, M. J., and W. Schlenker, 2013: Identifying supply and demand elasticities of agricultural
5 commodities: Implications for the US ethanol mandate. *Am. Econ. Rev.*,
6 doi:10.1257/aer.103.6.2265.
- 7 Robinson, B. E., M. B. Holland, and L. Naughton-Treves, 2014: Does secure land tenure save forests?
8 A meta-analysis of the relationship between land tenure and tropical deforestation. *Glob.*
9 *Environ. Chang.*, **29**, 281–293, doi:10.1016/J.GLOENVCHA.2013.05.012.
10 <https://www.sciencedirect.com/science/article/pii/S0959378013000976> (Accessed May 23,
11 2018).
- 12 Robledo-Abad, C., and Coauthors, 2017: Bioenergy production and sustainable development: science
13 base for policymaking remains limited. *GCB Bioenergy*, **9**, 541–556, doi:10.1111/gcbb.12338.
- 14 Rocca, M. E., P. M. Brown, L. H. MacDonald, and C. M. Carrico, 2014: Climate change impacts on
15 fire regimes and key ecosystem services in Rocky Mountain forests. *For. Ecol. Manage.*, **327**,
16 290–305, doi:10.1016/j.foreco.2014.04.005.
- 17 Rochecouste, J.-F., P. Dargusch, D. Cameron, and C. Smith, 2015: An analysis of the socio-economic
18 factors influencing the adoption of conservation agriculture as a climate change mitigation
19 activity in Australian dryland grain production. *Agric. Syst.*, **135**, 20–30,
20 doi:http://dx.doi.org/10.1016/j.agsy.2014.12.002.
21 <http://www.sciencedirect.com/science/article/pii/S0308521X1400170X>.
- 22 Rocheleau, D., and D. Edmunds, 1997: Women, men and trees: Gender, power and property in forest
23 and agrarian landscapes. *World Dev.*, **25**, 1351–1371.
- 24 Rockström, Johan Steffen, W., and Coauthors, 2009: A safe operating space for humanity. *Nature*,
25 doi:10.1038/461472a.
- 26 Rockström, J., and Coauthors, 2009: A safe operating space for humanity. *Nature*, **461**, 472–475,
27 doi:10.1038/461472a.
- 28 —, and Coauthors, 2017: Sustainable intensification of agriculture for human prosperity and global
29 sustainability. *Ambio*, **46**, 4–17.
- 30 Rocle, N., and D. Salles, 2018: “Pioneers but not guinea pigs”: experimenting with climate change
31 adaptation in French coastal areas. *Policy Sci.*, doi:10.1007/s11077-017-9279-z.
- 32 Rodell, M., I. Velicogna, and J. S. Famiglietti, 2009: Satellite-based estimates of groundwater
33 depletion in India. *Nature*, **460**, 999–1002, doi:10.1038/nature08238.
34 <http://www.ncbi.nlm.nih.gov/pubmed/19675570> (Accessed May 18, 2018).
- 35 Rodríguez-Morales, J. E., 2018: Convergence, conflict and the historical transition of bioenergy for
36 transport in Brazil: The political economy of governance and institutional change. *Energy Res.*
37 *Soc. Sci.*, **44**, 324–335, doi:10.1016/j.erss.2018.05.031.
38 <https://linkinghub.elsevier.com/retrieve/pii/S2214629618305346> (Accessed April 13, 2019).
- 39 Rodríguez-Takeuchi, L., and K. S. Imai, 2013: Food price surges and poverty in urban colombia: New
40 evidence from household survey data. *Food Policy*, doi:10.1016/j.foodpol.2013.09.017.
- 41 Rodríguez-Ward, D., A. M. Larson, and H. G. Ruesta, 2018: Top-down, Bottom-up and Sideways:
42 The Multilayered Complexities of Multi-level Actors Shaping Forest Governance and REDD+
43 Arrangements in Madre de Dios, Peru. *Environ. Manage.*, **62**, 98–116, doi:10.1007/s00267-017-
44 0982-5. <http://link.springer.com/10.1007/s00267-017-0982-5> (Accessed October 15, 2018).
- 45 Rodríguez, J., T. D. Beard Jr, E. Bennett, G. Cumming, S. Cork, J. Agard, A. Dobson, and G.
46 Peterson, 2006: Trade-offs across space, time, and ecosystem services. *Ecol. Soc.*, **11**.
- 47 Rodríguez Morales, J. E., and F. Rodríguez López, 2017: The political economy of bioenergy in the

- 1 United States: A historical perspective based on scenarios of conflict and convergence. *Energy*
2 *Res. Soc. Sci.*, doi:10.1016/j.erss.2017.03.002.
- 3 Roelich, K., and J. Gieseckam, 2019: Decision making under uncertainty in climate change mitigation:
4 introducing multiple actor motivations, agency and influence. *Clim. Policy*, **19**, 175–188,
5 doi:10.1080/14693062.2018.1479238.
- 6 Rogelj, J., and Coauthors, 2016: Paris Agreement climate proposals need a boost to keep warming
7 well below 2 °C. *Nature*, **534**, 631–639, doi:10.1038/nature18307.
8 <http://www.nature.com/doi/finder/10.1038/nature18307>.
- 9 Rogelj, J., and Coauthors, 2018: Mitigation pathways compatible with 1.5°C in the context of
10 sustainable development. *Global Warming of 1.5 °C an IPCC special report on the impacts of*
11 *global warming of 1.5 °C above pre-industrial levels and related global greenhouse gas*
12 *emission pathways, in the context of strengthening the global response to the threat of climate*
13 *change* <http://www.ipcc.ch/report/sr15/>.
- 14 Rogge, K. S., and K. Reichardt, 2016: Policy mixes for sustainability transitions: An extended
15 concept and framework for analysis. *Res. Policy*, doi:10.1016/j.respol.2016.04.004.
- 16 Romijn, H. A., 2011: Land clearing and greenhouse gas emissions from Jatropha biofuels on African
17 Miombo Woodlands. *Energy Policy*, **39**, 5751–5762, doi:10.1016/J.ENPOL.2010.07.041.
18 <https://www.sciencedirect.com/science/article/pii/S0301421510005719> (Accessed April 9,
19 2019).
- 20 Roncoli, C., C. Jost, C. Perez, K. Moore, A. Ballo, S. Cissé, and K. Ouattara, 2007: Carbon
21 sequestration from common property resources: Lessons from community-based sustainable
22 pasture management in north-central Mali. *Agric. Syst.*, **94**, 97–109,
23 doi:10.1016/j.agsy.2005.10.010.
24 <http://linkinghub.elsevier.com/retrieve/pii/S0308521X06001132> (Accessed November 3, 2018).
- 25 Rosenthal, J., A. Quinn, A. P. Grieshop, A. Pillarisetti, and R. I. Glass, 2017: Clean cooking and the
26 SDGs: Integrated analytical approaches to guide energy interventions for health and
27 environment goals ☆ , ☆☆. doi:10.1016/j.esd.2017.11.003.
28 <https://doi.org/10.1016/j.esd.2017.11.003> (Accessed April 15, 2019).
- 29 Rosenzweig, C., and Coauthors, 2014: Assessing agricultural risks of climate change in the 21st
30 century in a global gridded crop model intercomparison. *Proc. Natl. Acad. Sci.*,
31 doi:10.1073/pnas.1222463110.
- 32 Rosillo Callé, F., and F. X. Johnson, 2010a: *Food versus fuel : an informed introduction to biofuels*.
33 Zed Books, 217 pp. [http://www.diva-](http://www.diva-portal.org.ezp.sub.su.se/smash/record.jsf?pid=diva2%3A501125&dswid=-6156)
34 [portal.org.ezp.sub.su.se/smash/record.jsf?pid=diva2%3A501125&dswid=-6156](http://www.diva-portal.org.ezp.sub.su.se/smash/record.jsf?pid=diva2%3A501125&dswid=-6156) (Accessed April
35 8, 2019).
- 36 —, and —, 2010b: *Food versus fuel : an informed introduction to biofuels*. Zed Books, 217 pp.
- 37 Rosin, C., 2013: Food security and the justification of productivism in New Zealand. *J. Rural Stud.*,
38 **29**, 50–58, doi:10.1016/j.jrurstud.2012.01.015.
- 39 Roudier, P., B. Muller, P. Aquino, C. Roncoli, M. A. Soumaré, L. Batté, and B. Sultan, 2014: Climate
40 Risk Management The role of climate forecasts in smallholder agriculture: Lessons from
41 participatory research in two communities in Senegal. *Clim. Risk Manag.*,
42 doi:10.1016/j.crm.2014.02.001.
- 43 Rouillard, J., D. Benson, A. K. Gain, and C. Giupponi, 2017: Governing for the Nexus: Empirical,
44 Theoretical and Normative Dimensions. *Water-energy-food Nexus Princ. Pract.*,.
- 45 Rouillard, J. J., K. V. Heal, T. Ball, and A. D. Reeves, 2013: Policy integration for adaptive water
46 governance: Learning from Scotland’s experience. *Environ. Sci. Policy*,
47 doi:10.1016/j.envsci.2013.07.003.

- 1 Roy, J., and Coauthors, 2018: Sustainable Development , Poverty Eradication and Reducing
2 Inequalities. *Global Warming of 1.5 °C an IPCC special report on the impacts of global*
3 *warming of 1.5 °C above pre-industrial levels and related global greenhouse gas emission*
4 *pathways, in the context of strengthening the global response to the threat of climate change*
5 <http://www.ipcc.ch/report/sr15/>.
- 6 Rozin, P., S. Scott, M. Dingley, J. K. Urbanek, H. Jiang, and M. Kaltenbach, 2011: Nudge to nobesity
7 I: Minor changes in accessibility decrease food intake. *Judgm. Decis. Mak.*, **6**, 323–332.
8 [https://www.scopus.com/inward/record.uri?eid=2-s2.0-](https://www.scopus.com/inward/record.uri?eid=2-s2.0-79960040698&partnerID=40&md5=4044400d8018add3e8a59f5558f9c460)
9 [79960040698&partnerID=40&md5=4044400d8018add3e8a59f5558f9c460](https://www.scopus.com/inward/record.uri?eid=2-s2.0-79960040698&partnerID=40&md5=4044400d8018add3e8a59f5558f9c460).
- 10 Rudolph, D. L., J. F. Devlin, and L. Bekeris, 2015: Challenges and a strategy for agricultural BMP
11 monitoring and remediation of nitrate contamination in unconsolidated aquifers. *Groundw.*
12 *Monit. Remediat.*, doi:10.1111/gwmr.12103.
- 13 Ruiz-Mercado, I., and O. Masera, 2015a: Patterns of Stove Use in the Context of Fuel-Device
14 Stacking: Rationale and Implications. *Ecohealth*, **12**, 42–56, doi:10.1007/s10393-015-1009-4.
15 <https://link-springer-com.ezp.sub.su.se/content/pdf/10.1007%2Fs10393-015-1009-4.pdf>
16 (Accessed April 14, 2019).
- 17 —, and —, 2015b: Patterns of Stove Use in the Context of Fuel-Device Stacking: Rationale and
18 Implications. *Ecohealth*, **12**, 42–56, doi:10.1007/s10393-015-1009-4.
- 19 Rulli, M. C., A. Savioli, and P. D’Odorico, 2013: Global land and water grabbing. *Proc. Natl. Acad.*
20 *Sci.*, **110**, 892–897, doi:10.1073/PNAS.1213163110.
21 <https://www.pnas.org/content/110/3/892.short> (Accessed April 12, 2019).
- 22 —, D. Bellomi, A. Cazzoli, G. De Carolis, and P. D’Odorico, 2016: The water-land-food nexus of
23 first-generation biofuels. *Sci. Rep.*, doi:10.1038/srep22521.
- 24 Ryan, S. J., A. McNally, L. R. Johnson, E. A. Mordecai, T. Ben-Horin, K. Paaijmans, and K. D.
25 Lafferty, 2015: Mapping Physiological Suitability Limits for Malaria in Africa Under Climate
26 Change. *Vector-Borne Zoonotic Dis.*, **15**, 718–725, doi:10.1089/vbz.2015.1822.
27 <http://online.liebertpub.com/doi/10.1089/vbz.2015.1822> (Accessed December 29, 2017).
- 28 Safriel, U., 2017: Land degradation neutrality (LDN) in drylands and beyond – where has it come
29 from and where does it go. *Silva Fenn.*, doi:10.14214/sf.1650.
- 30 —, and Z. Adeel, 2008: Development paths of drylands: thresholds and sustainability. *Sustain. Sci.*,
31 **3**, 117–123, doi:10.1007/s11625-007-0038-5. [http://link.springer.com/10.1007/s11625-007-](http://link.springer.com/10.1007/s11625-007-0038-5)
32 [0038-5](http://link.springer.com/10.1007/s11625-007-0038-5) (Accessed April 8, 2019).
- 33 Salehyan, I., and C. S. Hendrix, 2014: Climate shocks and political violence. *Glob. Environ. Chang.*,
34 doi:10.1016/j.gloenvcha.2014.07.007.
- 35 Saluja, N and Singh, S., 2018: Coal-fired power plants set to get renewed push. *Economic Times*, June
36 27 [https://economictimes.indiatimes.com/industry/energy/power/coal-fired-power-plants-set-to-](https://economictimes.indiatimes.com/industry/energy/power/coal-fired-power-plants-set-to-get-renewed-push/articleshow/64769464.cms)
37 [get-renewed-push/articleshow/64769464.cms](https://economictimes.indiatimes.com/industry/energy/power/coal-fired-power-plants-set-to-get-renewed-push/articleshow/64769464.cms).
- 38 Salvati, L., and M. Carlucci, 2014: Zero Net Land Degradation in Italy: The role of socioeconomic
39 and agro-forest factors. *J. Environ. Manage.*, **145**, 299–306, doi:10.1016/j.jenvman.2014.07.006.
- 40 Salzman, J., G. Bennett, N. Carroll, A. Goldstein, and M. Jenkins, 2018: The global status and trends
41 of Payments for Ecosystem Services. *Nat. Sustain.*, doi:10.1038/s41893-018-0033-0.
- 42 Samaddar, S., M. Yokomatsu, F. Dayour, M. Oteng-Ababio, T. Dzivenu, M. Adams, and H. Ishikawa,
43 2015: Evaluating Effective Public Participation in Disaster Management and Climate Change
44 Adaptation: Insights From Northern Ghana Through a User-Based Approach. *Risk, Hazards*
45 *Cris. Public Policy*, **6**, 117–143, doi:10.1002/rhc3.12075.
46 <http://doi.wiley.com/10.1002/rhc3.12075> (Accessed May 22, 2018).
- 47 Samanta, A., S. Ganguly, H. Hashimoto, S. Devadiga, E. Vermote, Y. Knyazikhin, R. R. Nemani, and
48 R. B. Myneni, 2010: Amazon forests did not green-up during the 2005 drought. *Geophys. Res.*

- 1 *Lett.*, **37**.
- 2 Samuwai, J., and J. . Hills, 2018: Assessing Climate Finance Readiness in the Asia-Pacific Region. 1–
3 18, doi:10.3390/su10041192.
- 4 Sánchez, B., A. Iglesias, A. McVittie, J. Álvaro-Fuentes, J. Ingram, J. Mills, J. P. Lesschen, and P. J.
5 Kuikman, 2016: Management of agricultural soils for greenhouse gas mitigation: Learning from
6 a case study in NE Spain. *J. Environ. Manage.*, **170**, 37–49,
7 doi:http://dx.doi.org/10.1016/j.jenvman.2016.01.003.
8 http://www.sciencedirect.com/science/article/pii/S0301479716300032.
- 9 Sánchez, J. M. T., and R. C. Maseda, 2016: Forcing and avoiding change. Exploring change and
10 continuity in local land-use planning in Galicia (Northwest of Spain) and The Netherlands. *Land*
11 *use policy*, **50**, 74–82, doi:10.1016/J.LANDUSEPOL.2015.09.006.
12 https://www.sciencedirect.com/science/article/pii/S0264837715002768 (Accessed May 23,
13 2018).
- 14 Sanderson, T., G. Hertzler, T. Capon, and P. Hayman, 2016: A real options analysis of Australian
15 wheat production under climate change. *Aust. J. Agric. Resour. Econ.*, **60**, 79–96,
16 doi:10.1111/1467-8489.12104.
- 17 Sandifer, P. A., A. E. Sutton-Grier, and B. P. Ward, 2015: Exploring connections among nature,
18 biodiversity, ecosystem services, and human health and well-being: Opportunities to enhance
19 health and biodiversity conservation. *Ecosyst. Serv.*, doi:10.1016/j.ecoser.2014.12.007.
- 20 Sandstrom, S., and S. Juhola, 2017: Continue to blame it on the rain? Conceptualization of drought
21 and failure of food systems in the Greater Horn of Africa. *Environ. Hazards*, **16**, 71–91.
- 22 Santos, M. J., S. C. Dekker, V. Daioglou, M. C. Braakhekke, and D. P. van Vuuren, 2017: Modeling
23 the Effects of Future Growing Demand for Charcoal in the Tropics. *Front. Environ. Sci.*, **5**,
24 doi:10.3389/fenvs.2017.00028.
25 http://journal.frontiersin.org/article/10.3389/fenvs.2017.00028/full (Accessed April 5, 2019).
- 26 Sanz, M. J., and Coauthors, 2017: *Sustainable Land Management contribution to successful land-*
27 *based climate change adaptation and mitigation. A Report of the Science-Policy Interface.*
28 Bonn, Germany, https://www.unccd.int/sites/default/files/documents/2017-
29 09/UNCCD_Report_SLM.pdf (Accessed March 29, 2019).
- 30 Sarap, K., T. K. Sarangi, and J. Naik, 2013: Implementation of Forest Rights Act 2006 in Odisha:
31 process, constraints and outcome. *Econ. Polit. Wkly.*, 61–67.
- 32 Sarzynski, A., 2015: Public participation, civic capacity, and climate change adaptation in cities.
33 *Urban Clim.*, **14**, 52–67, doi:10.1016/J.UCLIM.2015.08.002.
34 https://www.sciencedirect.com/science/article/pii/S2212095515300158 (Accessed May 16,
35 2018).
- 36 Sassi, F., and Coauthors, 2018: Equity impacts of price policies to promote healthy behaviours. *The*
37 *Lancet*.
- 38 Satterthwaite, D., 2007: *Climate Change and Urbanization: Effects and Implications for Urban*
39 *Governance*. http://www.un.org/esa/population/meetings/EGM_PopDist/P16_Satterthwaite.pdf
40 (Accessed November 3, 2018).
- 41 ———, D. Archer, S. Colenbrander, D. Dodman, J. Hardoy, and S. Patel, 2018: *Responding to climate*
42 *change in cities and in their informal settlements and economies*. https://citiesipcc.org/wp-
43 content/uploads/2018/03/Informality-background-paper-for-IPCC-Cities.pdf (Accessed
44 November 3, 2018).
- 45 Sauerwald, S., and M. W. Peng, 2013: Informal institutions, shareholder coalitions, and principal-
46 principal conflicts. *Asia Pacific J. Manag.*, **30**, 853–870, doi:10.1007/s10490-012-9312-x.
- 47 Savaresi, A., 2016: Journal of Energy & Natural Resources Law The Paris Agreement: a new
48 beginning? The Paris Agreement: a new beginning? *J. Energy Nat. Resour. Law*, **34**, 16–26,

- 1 doi:10.1080/02646811.2016.1133983.
2 <http://www.tandfonline.com/action/journalInformation?journalCode=nrnl20> (Accessed May 22,
3 2018).
- 4 Sayer, J. Margules, C., Boedhihartono, A., 2015: The Role of Citizen Science in Landscape and
5 Seascape Approaches to Integrating Conservation and Development. *Land*, **4**, 1200–1212.
- 6 Scarano, F. R., 2017: Ecosystem-based adaptation to climate change: concept, scalability and a role
7 for conservation science. *Perspect. Ecol. Conserv.*, **15**, 65–73, doi:10.1016/j.pecon.2017.05.003.
- 8 Scarlat, N., and J.-F. Dallemand, 2011: Recent developments of biofuels/bioenergy sustainability
9 certification: A global overview. *Energy Policy*, **39**, 1630–1646,
10 doi:10.1016/J.ENPOL.2010.12.039. [https://www.sciencedirect-](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0301421510009390)
11 [com.ezp.sub.su.se/science/article/pii/S0301421510009390](https://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0301421510009390) (Accessed May 23, 2018).
- 12 Schalatek, L., and S. Nakhouda, 2013: The Green Climate Fund. *Clim. Financ. Fundam.*, **11**, 1–4.
- 13 Scheffran, J., E. Marmer, and P. Sow, 2012: Migration as a contribution to resilience and innovation
14 in climate adaptation: Social networks and co-development in Northwest Africa. *Appl. Geogr.*,
15 doi:10.1016/j.apgeog.2011.10.002.
- 16 Scherr, S. J., S. Shames, and R. Friedman, 2012: From climate-smart agriculture to climate-smart
17 landscapes. *Agric. Food Secur.*, **1**, 12, doi:10.1186/2048-7010-1-12.
18 <http://agricultureandfoodsecurity.biomedcentral.com/articles/10.1186/2048-7010-1-12>
19 (Accessed December 21, 2017).
- 20 Schick, A., C. Sandig, A. Krause, P. R. Hobson, S. Porembski, and P. L. Ibisch, 2018: People-
21 Centered and Ecosystem-Based Knowledge Co-Production to Promote Proactive Biodiversity
22 Conservation and Sustainable Development in Namibia. *Environ. Manage.*, **62**, 858–876,
23 doi:10.1007/s00267-018-1093-7. <http://link.springer.com/10.1007/s00267-018-1093-7>
24 (Accessed April 8, 2019).
- 25 Schlager, E., and E. Ostrom, 1992: Property-Rights Regimes and Natural Resources: A Conceptual
26 Analysis. *Land Econ.*, **68**, 249, doi:10.2307/3146375.
27 <http://www.jstor.org/stable/3146375?origin=crossref> (Accessed May 22, 2018).
- 28 Schleussner, C. F., and Coauthors, 2016: Differential climate impacts for policy-relevant limits to
29 global warming: The case of 1.5 °c and 2 °c. *Earth Syst. Dyn.*, **7**, 327–351, doi:10.5194/esd-7-
30 327-2016.
- 31 Schmalensee, R., and R. N. Stavins, 2017: Lessons learned from three decades of experience with cap
32 and trade. *Rev. Environ. Econ. Policy*, **11**, 59–79, doi:10.1093/reep/rew017.
- 33 Schmidhuber, J., and F. N. Tubiello, 2007: Global food security under climate change. *Proc. Natl.*
34 *Acad. Sci. U. S. A.*, **104**, 19703–19708, doi:10.1073/pnas.0701976104.
35 <http://www.ncbi.nlm.nih.gov/pubmed/18077404> (Accessed May 18, 2018).
- 36 Schmitz, C., A. Biewald, H. Lotze-Campen, A. Popp, J. P. Dietrich, B. Bodirsky, M. Krause, and I.
37 Weindl, 2012: Trading more food: Implications for land use, greenhouse gas emissions, and the
38 food system. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2011.09.013.
- 39 Schmitz, O. J., and Coauthors, 2015: Conserving Biodiversity: Practical Guidance about Climate
40 Change Adaptation Approaches in Support of Land-use Planning. *Source Nat. Areas J.*, **35**, 190–
41 203, doi:10.3375/043.035.0120. [https://depts.washington.edu/landecol/PDFS/Schmitz](https://depts.washington.edu/landecol/PDFS/Schmitz_et) et
42 al.2015.pdf (Accessed February 25, 2019).
- 43 Schneider, L., and S. La Hoz Theuer, 2019: Environmental integrity of international carbon market
44 mechanisms under the Paris Agreement. *Clim. Policy*, **19**, 386–400,
45 doi:10.1080/14693062.2018.1521332.
46 <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1521332> (Accessed March 17,
47 2019).
- 48 Schröder, P., and Coauthors, 2018: Intensify production, transform biomass to energy and novel

- 1 goods and protect soils in Europe—A vision how to mobilize marginal lands. *Sci. Total*
2 *Environ.*, doi:10.1016/j.scitotenv.2017.10.209.
- 3 Schultz, L., C. Folke, H. Österblom, and P. Olsson, 2015: Adaptive governance, ecosystem
4 management, and natural capital. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1406493112.
- 5 Schut, M., N. C. Soares, G. Van De Ven, and M. Slingerland, 2013: Multi-actor governance of
6 sustainable biofuels in developing countries: The case of Mozambique.
7 doi:10.1016/j.enpol.2013.09.007. <http://dx.doi.org/10.1016/j.enpol.2013.09.007> (Accessed April
8 12, 2019).
- 9 De Schutter, O., 2011: How not to think of land-grabbing: three critiques of large-scale investments in
10 farmland. *J. Peasant Stud.*, **38**, 249–279, doi:10.1080/03066150.2011.559008.
11 <http://www.tandfonline.com/action/journalInformation?journalCode=fjps20> (Accessed
12 November 2, 2018).
- 13 Schuur, E. A. G., and Coauthors, 2015: Climate change and the permafrost carbon feedback. *Nature*,
14 **520**, 171–179, doi:10.1038/nature14338. <http://www.nature.com/articles/nature14338> (Accessed
15 May 24, 2018).
- 16 Schwartz, N. B., M. Uriarte, R. DeFries, V. H. Gutierrez-Velez, and M. A. Pinedo-Vasquez, 2017:
17 Land-use dynamics influence estimates of carbon sequestration potential in tropical second-
18 growth forest. *Environ. Res. Lett.*, **12**, 074023, doi:10.1088/1748-9326/aa708b.
19 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=7/a=074023?key=crossref.de820ae6e07cc2cc7d048a9a07b54c13)
20 [9326/12/i=7/a=074023?key=crossref.de820ae6e07cc2cc7d048a9a07b54c13](http://stacks.iop.org/1748-9326/12/i=7/a=074023?key=crossref.de820ae6e07cc2cc7d048a9a07b54c13) (Accessed May 22,
21 2018).
- 22 Schwilch, G., and Coauthors, 2011: Experiences in monitoring and assessment of sustainable land
23 management. *L. Degrad. Dev.*, **22**, 214–225, doi:10.1002/ldr.1040.
24 <http://doi.wiley.com/10.1002/ldr.1040> (Accessed April 9, 2019).
- 25 Scott, C. A., S. A. Pierce, M. J. Pasqualetti, A. L. Jones, B. E. Montz, and J. H. Hoover, 2011: Policy
26 and institutional dimensions of the water-energy nexus. *Energy Policy*, **39**, 6622–6630,
27 doi:10.1016/j.enpol.2011.08.013.
- 28 Scott, D., C. M. Hall, and S. Gössling, 2016: A report on the Paris Climate Change Agreement and its
29 implications for tourism: Why we will always have Paris. *J. Sustain. Tour.*, **24**, 933–948.
- 30 Seager, J., 2014: Disasters are gendered: What’s new? *Reducing disaster: Early warning systems For*
31 *climate change*, A. Singh and Z. Zommers, Eds., Springer, Dordrecht, 265–281.
- 32 Seaman, J. A., G. E. Sawdon, J. Acidri, and C. Petty, 2014: The household economy approach.
33 managing the impact of climate change on poverty and food security in developing countries.
34 *Clim. Risk Manag.*, doi:10.1016/j.crm.2014.10.001.
- 35 Selvaraju, R., 2011: Climate risk assessment and management in agriculture. *Build. Resil. Adapt. to*
36 *Clim. Chang. Agric. Sect.*, doi:doi:10.1103/PhysRevD.64.104020.
- 37 Selvaraju, R., 2012: Climate risk assessment and management in agriculture. *Building Resilience for*
38 *Adaptation to Climate change in the Agriculture Sector, Proceedings of a Joint FAO/OECD*
39 *Workshop, April 23-24, 2012*, V. Meybeck, A., Lankoski, J., Redfern S., Azzu, N., Gitz, Ed., 71–
40 90
41 [https://www.researchgate.net/publication/303539375_Climate_risk_assessment_and_manageme](https://www.researchgate.net/publication/303539375_Climate_risk_assessment_and_management_in_agriculture)
42 [nt_in_agriculture](https://www.researchgate.net/publication/303539375_Climate_risk_assessment_and_management_in_agriculture) (Accessed February 25, 2019).
- 43 Seng, D. C., 2013: Improving the Governance Context and Framework Conditions of Natural Hazard
44 Early Warning Systems. *J. Integr. Disaster Risk Manag.*, doi:10.5595/idrim.2012.0020.
- 45 Serrao-Neumann, S., B. P. Harman, and D. Low Choy, 2013: The Role of Anticipatory Governance in
46 Local Climate Adaptation: Observations from Australia. *Plan. Pract. Res.*,
47 doi:10.1080/02697459.2013.795788.
- 48 —, F. Crick, B. Harman, G. Schuch, and D. L. Choy, 2015a: Maximising synergies between

- 1 disaster risk reduction and climate change adaptation: Potential enablers for improved planning
2 outcomes. *Environ. Sci. Policy*, **50**, 46–61, doi:10.1016/j.envsci.2015.01.017.
- 3 —, B. Harman, A. Leitch, and D. Low Choy, 2015b: Public engagement and climate adaptation:
4 insights from three local governments in Australia. *J. Environ. Plan. Manag.*, **58**, 1196–1216,
5 doi:10.1080/09640568.2014.920306.
- 6 Seto, K. C., 2011: Exploring the dynamics of migration to mega-delta cities in Asia and Africa:
7 Contemporary drivers and future scenarios. *Glob. Environ. Chang.*,
8 doi:10.1016/j.gloenvcha.2011.08.005.
- 9 Shaaban, M., L. Van Zwieten, S. Bashir, and et al, 2018: A concise review of biochar application to
10 agricultural soils to improve soil conditions and fight pollution. *J. Environ. Manage.*, **228**, 429–
11 440, doi:10.1016/J.JENVMAN.2018.09.006. [https://www.sciencedirect-](https://www.sciencedirect.com/ezp.sub.su.se/science/article/pii/S0301479718309976?via%3Dihub)
12 [com/ezp.sub.su.se/science/article/pii/S0301479718309976?via%3Dihub](https://www.sciencedirect.com/ezp.sub.su.se/science/article/pii/S0301479718309976?via%3Dihub) (Accessed April 15,
13 2019).
- 14 Sharples, J. J., and Coauthors, 2016a: Natural hazards in Australia: extreme bushfire. *Clim. Change*,
15 **139**, 85–99, doi:10.1007/s10584-016-1811-1. <http://dx.doi.org/10.1007/s10584-016-1811-1>.
- 16 —, and Coauthors, 2016b: Natural hazards in Australia: extreme bushfire. *Clim. Change*, **139**, 85–
17 99, doi:10.1007/s10584-016-1811-1.
- 18 Sheil, D., M. Boissière, and G. Beaudoin, 2015: Unseen sentinels: local monitoring and control in
19 conservation’s blind spots. *Ecol. Soc.*, **20**, art39, doi:10.5751/ES-07625-200239.
20 <http://www.ecologyandsociety.org/vol20/iss2/art39/> (Accessed April 9, 2019).
- 21 Sheng, J., X. Han, H. Zhou, and Z. Miao, 2016: Effects of corruption on performance: Evidence from
22 the UN-REDD Programme. *Land use policy*, **59**, 344–350,
23 doi:10.1016/j.landusepol.2016.09.014.
24 <https://linkinghub.elsevier.com/retrieve/pii/S026483771630134X> (Accessed November 3, 2018).
- 25 Shiferaw, B., K. Tesfaye, M. Kassie, T. Abate, B. M. Prasanna, and A. Menkir, 2014: Managing
26 vulnerability to drought and enhancing livelihood resilience in sub-Saharan Africa:
27 Technological, institutional and policy options. *Weather Clim. Extrem.*, **3**, 67–79,
28 doi:10.1016/j.wace.2014.04.004. <http://libcatalog.cimmyt.org/Download/cis/98992.pdf>
29 (Accessed May 16, 2018).
- 30 Shindell, D., and Coauthors, 2012: Simultaneously mitigating near-term climate change and
31 improving human health and food security. *Science*, **335**, 183–189,
32 doi:10.1126/science.1210026. <http://www.ncbi.nlm.nih.gov/pubmed/22246768> (Accessed
33 November 16, 2018).
- 34 Shogren, J. ., and L. . Taylor, 2008: On behavioural-environmental economics. *Rev. Environ. Econ.*
35 *Policy*, **2**.
- 36 Shreve, C. M., and I. Kelman, 2014: Does mitigation save? Reviewing cost-benefit analyses of
37 disaster risk reduction. doi:10.1016/j.ijdr.2014.08.004.
38 <https://pdfs.semanticscholar.org/2b43/95d2e6471b5bd3743b66eec6ae28b68a474c.pdf> (Accessed
39 April 3, 2019).
- 40 Shue, H., 2018a: Mitigation gambles: uncertainty, urgency and the last gamble possible. *Philos.*
41 *Trans. R. Soc. A Math. Eng. Sci.*, **376**, 20170105, doi:10.1098/rsta.2017.0105.
42 <http://rsta.royalsocietypublishing.org/lookup/doi/10.1098/rsta.2017.0105> (Accessed March 30,
43 2019).
- 44 —, 2018b: Mitigation gambles: uncertainty, urgency and the last gamble possible. *Philos. Trans. R.*
45 *Soc. A Math. Eng. Sci.*, **376**, 20170105, doi:10.1098/rsta.2017.0105.
- 46 Shvidenko, A. Z., D. G. Shchepashchenko, E. A. Vaganov, A. I. Sukhinin, S. S. Maksyutov, I.
47 McCallum, and I. P. Lakyda, 2012: Impact of wildfire in Russia between 1998–2010 on
48 ecosystems and the global carbon budget. *Dokl. Earth Sci.*, **441**, 1678–1682,

- 1 doi:10.1134/s1028334x11120075.
- 2 Siahaya, M. E., T. R. Hutaeruk, H. S. E. S. Aponno, J. W. Hatulesila, and A. B. Mardhanie, 2016:
3 Traditional ecological knowledge on shifting cultivation and forest management in East Borneo,
4 Indonesia. *Int. J. Biodivers. Sci. Ecosyst. Serv. Manag.*, **12**, 14–23,
5 doi:10.1080/21513732.2016.1169559.
6 <https://www.tandfonline.com/doi/full/10.1080/21513732.2016.1169559> (Accessed April 8,
7 2019).
- 8 Siddig, E. F. A., K. El - Harizi, and B. Prato, 2007: Managing Conflict Over Natural Resources in
9 Greater Kordofan, Sudan: Some Recurrent Patterns and Governance Implications. *Int. Food*
10 *Policy Res. Cent.*.
- 11 Siebert, A., 2016: Analysis of the future potential of index insurance in the West African Sahel using
12 CMIP5 GCM results. *Clim. Change*, **134**, 15–28, doi:10.1007/s10584-015-1508-x.
13 <http://link.springer.com/10.1007/s10584-015-1508-x> (Accessed May 16, 2018).
- 14 Siegmeier, J., and Coauthors, 2018: The fiscal benefits of stringent climate change mitigation : an
15 overview. **3062**, doi:10.1080/14693062.2017.1400943.
- 16 Sietz, D., L. Fleskens, and L. C. Stringer, 2017: Learning from Non-Linear Ecosystem Dynamics Is
17 Vital for Achieving Land Degradation Neutrality. *L. Degrad. Dev.*, **28**, 2308–2314,
18 doi:10.1002/ldr.2732.
- 19 Sigurdsson, J. H., L. A. Walls, and J. L. Quigley, 2001: Bayesian belief nets for managing expert
20 judgement and modelling reliability. *Qual. Reliab. Eng. Int.*, **17**, 181–190, doi:10.1002/qre.410.
- 21 Silva-Olaya, A. M., C. E. P. Cerri, S. Williams, C. C. Cerri, C. A. Davies, and K. Paustian, 2017:
22 Modelling SOC response to land use change and management practices in sugarcane cultivation
23 in South-Central Brazil. *Plant Soil*, doi:10.1007/s11104-016-3030-y.
- 24 Silva, R. A., and Coauthors, 2013: Global premature mortality due to anthropogenic outdoor air
25 pollution and the contribution of past climate change. *Environ. Res. Lett.*, doi:10.1088/1748-
26 9326/8/3/034005.
- 27 —, and Coauthors, 2016: The effect of future ambient air pollution on human premature mortality
28 to 2100 using output from the ACCMIP model ensemble. *Atmos. Chem. Phys.*, doi:10.5194/acp-
29 16-9847-2016.
- 30 Silvano, R. A. M., and J. Valbo-Jørgensen, 2008: Beyond fishermen’s tales: contributions of fishers’
31 local ecological knowledge to fish ecology and fisheries management. *Environ. Dev. Sustain.*,
32 **10**, 657–675, doi:10.1007/s10668-008-9149-0. <http://link.springer.com/10.1007/s10668-008-9149-0>
33 (Accessed November 4, 2018).
- 34 Silveira, S., and F. X. Johnson, 2016: Navigating the transition to sustainable bioenergy in Sweden
35 and Brazil: Lessons learned in a European and International context. *Energy Res. Soc. Sci.*, **13**,
36 180–193, doi:10.1016/j.erss.2015.12.021.
37 <https://linkinghub.elsevier.com/retrieve/pii/S2214629615300839> (Accessed April 13, 2019).
- 38 Silvertown, J., 2009: A new dawn for citizen science. 467–471, doi:10.1016/j.tree.2009.03.017.
- 39 Simonet, G., J. Subervie, D. Ezzine-de-Blas, M. Cromberg, and A. E. Duchelle, 2019: Effectiveness
40 of a REDD+ Project in Reducing Deforestation in the Brazilian Amazon. *Am. J. Agric. Econ.*,
41 **101**, 211–229, doi:10.1093/ajae/aay028.
42 <https://academic.oup.com/ajae/article/101/1/211/5039934> (Accessed March 9, 2019).
- 43 Sindhu, S., V. Nehra, and S. Luthra, 2017: Investigation of feasibility study of solar farms deployment
44 using hybrid AHP-TOPSIS analysis: Case study of India. *Renew. Sustain. Energy Rev.*, **73**, 496–
45 511.
- 46 Singh, R., P. M. Reed, and K. Keller, 2015: Many-objective robust decision making for managing an
47 ecosystem with a deeply uncertain threshold response. *Ecol. Soc.*, **20**.

- 1 Slater, R., 2011: Cash transfers, social protection and poverty reduction. *Int. J. Soc. Welf.*,
2 doi:10.1111/j.1468-2397.2011.00801.x.
- 3 Smeets, E., F. X. Johnson, and G. Ballard-Tremeer, 2012a: Traditional and Improved Use of Biomass
4 for Energy in Africa. *Bioenergy for Sustainable Development in Africa*, Springer Netherlands,
5 Dordrecht, 3–12 http://www.springerlink.com/index/10.1007/978-94-007-2181-4_1 (Accessed
6 April 14, 2019).
- 7 ———, ———, and ———, 2012b: Traditional and Improved Use of Biomass for Energy in Africa.
8 *Bioenergy for Sustainable Development in Africa*, Springer Netherlands, Dordrecht, 3–12.
- 9 Smith, C. B., 2011: Adaptive management on the central Platte River. *J. Env. Manag.*, **92**, 1414–1419.
- 10 Smith, H. E., F. Eigenbrod, D. Kafumbata, M. D. Hudson, and K. Schreckenber, 2015: Criminals by
11 necessity: the risky life of charcoal transporters in Malawi. *For. Trees Livelihoods*, **24**, 259–274,
12 doi:10.1080/14728028.2015.1062808.
13 <http://www.tandfonline.com/doi/full/10.1080/14728028.2015.1062808> (Accessed November 16,
14 2018).
- 15 Smith, K. R., and Coauthors, 2014a: Millions Dead: How Do We Know and What Does It Mean?
16 Methods Used in the Comparative Risk Assessment of Household Air Pollution. *Annu. Rev.*
17 *Public Health*, **35**, 185–206, doi:10.1146/annurev-publhealth-032013-182356.
18 <http://www.annualreviews.org/doi/10.1146/annurev-publhealth-032013-182356> (Accessed May
19 23, 2018).
- 20 Smith, P., and J. E. Olesen, 2010: Synergies between the mitigation of, and adaptation to, climate
21 change in agriculture. *J. Agric. Sci.*, **148**, 543–552.
- 22 Smith, P., and Coauthors, 2007: Policy and technological constraints to implementation of greenhouse
23 gas mitigation options in agriculture. *Agric. Ecosyst. Environ.*, doi:10.1016/j.agee.2006.06.006.
- 24 Smith, P., and Coauthors, 2014b: Chapter 11 - Agriculture, forestry and other land use (AFOLU).
25 <http://pure.iiasa.ac.at/id/eprint/11115/> (Accessed April 25, 2019).
- 26 Smith, P., and Coauthors, 2016: Biophysical and economic limits to negative CO₂ emissions. *Nat.*
27 *Clim. Chang.*, doi:10.1038/nclimate2870.
- 28 Smucker, T. A., and E. E. Wangui, 2016: Gendered knowledge and adaptive practices: Differentiation
29 and change in Mwangi District, Tanzania. *Ambio*, **45**, 276–286, doi:10.1007/s13280-016-0828-
30 z. <http://link.springer.com/10.1007/s13280-016-0828-z> (Accessed April 2, 2019).
- 31 Soizic Le Saout, and A. S. L. R. Michael Hoffmann, Yichuan Shi, Adrian Hughes, Cyril
32 Bernard, Thomas M. Brooks, Bastian Bertzky, Stuart H.M. Butchart, Simon N. Stuart, Tim
33 Badman, 2013: Protected areas and effective biodiversity conservation. *Science (80-.)*, **342**,
34 803–805. <http://www.lerf.eco.br/img/publicacoes/Science-2013-AreasProtegidasMundo.pdf>
35 (Accessed May 17, 2018).
- 36 Sola, P., C. Ochieng, J. Yila, and M. Iiyama, 2016a: Links between energy access and food security in
37 sub Saharan Africa: an exploratory review. *Food Secur.*, **8**, 635–642, doi:10.1007/s12571-016-
38 0570-1. <http://link.springer.com/10.1007/s12571-016-0570-1> (Accessed April 2, 2019).
- 39 ———, ———, ———, and ———, 2016b: Links between energy access and food security in sub Saharan
40 Africa: an exploratory review. *Food Secur.*, **8**, 635–642, doi:10.1007/s12571-016-0570-1.
- 41 Solomon, D., and J. Lehmann, 2017: Socio-economic scenarios of low hanging fruits for developing
42 climate-smart biochar systems in Ethiopia: Biomass resource availability to sustainably improve
43 soil fertility, agricultural productivity and food and nutrition security.
44 <https://ecommons.cornell.edu/handle/1813/55321> (Accessed November 16, 2018).
- 45 Somanathan, E., R. Prabhakar, and B. S. Mehta, 2009: Decentralization for cost-effective
46 conservation. *Proc. Natl. Acad. Sci.*, **106**, 4143–4147.
- 47 Somanathan, E., and Coauthors, 2014: *15. National and Sub-national Policies and Institutions*. 1141-

- 1 1206 pp.
- 2 Song, X.-P., M. C. Hansen, S. V. Stehman, P. V. Potapov, A. Tyukavina, E. F. Vermote, and J. R.
3 Townshend, 2018: Global land change from 1982 to 2016. *Nature*, **560**, 639–643,
4 doi:10.1038/s41586-018-0411-9. <http://www.nature.com/articles/s41586-018-0411-9> (Accessed
5 April 14, 2019).
- 6 Sonnino, R., C. Lozano Torres, and S. Schneider, 2014: Reflexive governance for food security: The
7 example of school feeding in Brazil. *J. Rural Stud.*, **36**, 1–12,
8 doi:10.1016/j.jrurstud.2014.06.003.
- 9 Sorice, M. G., C. Josh Donlan, K. J. Boyle, W. Xu, and S. Gelcich, 2018: Scaling participation in
10 payments for ecosystem services programs. *PLoS One*, doi:10.1371/journal.pone.0192211.
- 11 Sorokin, A., A. Bryzhev, A. Stokov, A. Mirzabaev, T. Johnson, and S. V. Kiselev, 2015: The
12 Economics of land degradation in Russia. *Economics of Land Degradation and Improvement - A
13 Global Assessment for Sustainable Development*.
- 14 Sovacool, B. K., 2018: Bamboo Beating Bandits: Conflict, Inequality, and Vulnerability in the
15 Political Ecology of Climate Change Adaptation in Bangladesh. *World Dev.*, **102**, 183–194,
16 doi:10.1016/J.WORLDDEV.2017.10.014.
17 <https://www.sciencedirect.com/science/article/pii/S0305750X17303285> (Accessed May 23,
18 2018).
- 19 Sparrevik, M., C. Adam, V. Martinsen, and G. Cornelissen, 2015: Emissions of gases and particles
20 from charcoal/biochar production in rural areas using medium-sized traditional and improved
21 “retort” kilns. doi:10.1016/j.biombioe.2014.11.016.
22 www.sciencedirect.com<http://www.elsevier.com/locate/biombioe><http://dx.doi.org/10.1016/j.biombioe.2014.11.0160961-9534/> (Accessed April 15, 2019).
- 24 Sparrow, R., A. Suryahadi, and W. Widyanti, 2013: Social health insurance for the poor: Targeting
25 and impact of Indonesia’s Askeskin programme. *Soc. Sci. Med.*, **96**, 264–271.
- 26 Spence, A., W. Poortinga, and N. Pidgeon, 2012: The Psychological Distance of Climate Change.
27 *Risk Anal.*, **32**, 957–972, doi:10.1111/j.1539-6924.2011.01695.x.
- 28 Speranza, C. I., B. Kiteme, P. Ambenje, U. Wiesmann, and S. Makali, 2010: Indigenous knowledge
29 related to climate variability and change: Insights from droughts in semi-arid areas of former
30 Makueni District, Kenya. *Clim. Change*, **100**, 295–315, doi:10.1007/s10584-009-9713-0.
31 <http://link.springer.com/10.1007/s10584-009-9713-0> (Accessed January 2, 2018).
- 32 Spoon, J., 2014: Quantitative, qualitative, and collaborative methods: approaching indigenous
33 ecological knowledge heterogeneity. *Ecol. Soc.*, **19**, art33, doi:10.5751/ES-06549-190333.
34 <http://www.ecologyandsociety.org/vol19/iss3/art33/> (Accessed April 7, 2019).
- 35 Spracklen, D. V., S. R. Arnold, and C. M. Taylor, 2012: Observations of increased tropical rainfall
36 preceded by air passage over forests. *Nature*, **489**, 282.
- 37 Springmann, M., H. C. J. Godfray, M. Rayner, and P. Scarborough, 2016: Analysis and valuation of
38 the health and climate change cobenefits of dietary change. *Proc. Natl. Acad. Sci.*,
39 doi:10.1073/pnas.1523119113.
- 40 Stattman, S., A. Gupta, L. Partzsch, P. Oosterveer, S. L. Stattman, A. Gupta, L. Partzsch, and P.
41 Oosterveer, 2018a: Toward Sustainable Biofuels in the European Union? Lessons from a
42 Decade of Hybrid Biofuel Governance. *Sustainability*, **10**, 4111, doi:10.3390/su10114111.
43 <http://www.mdpi.com/2071-1050/10/11/4111> (Accessed April 8, 2019).
- 44 —, —, —, —, —, —, —, and —, 2018b: Toward Sustainable Biofuels in the
45 European Union? Lessons from a Decade of Hybrid Biofuel Governance. *Sustainability*, **10**,
46 4111, doi:10.3390/su10114111.
- 47 Stattman, S. L., A. Gupta, and L. Partzsch, *Biofuels in the European Union: Can Hybrid Governance
48 Promote Sustainability?* <https://ecpr.eu/Filestore/PaperProposal/4562fec9-d9e6-4806-876b->

- 1 17bfda3ff128.pdf (Accessed April 12, 2019).
- 2 Stavi, I., and R. Lal, 2015: Achieving Zero Net Land Degradation: Challenges and opportunities. *J.*
3 *Arid Environ.*, **112**, 44–51, doi:10.1016/j.jaridenv.2014.01.016.
- 4 Stavropoulou, M., R. Holmes, and N. Jones, 2017: Harnessing informal institutions to strengthen
5 social protection for the rural poor. *Glob. Food Sec.*, **12**, 73–79, doi:10.1016/j.gfs.2016.08.005.
- 6 Steffen, W., and Coauthors, 2015: Planetary boundaries: Guiding human development on a changing
7 planet. *Science (80-.)*, doi:10.1126/science.1259855.
- 8 Stephens, S., R. Bell, and J. Lawrence, 2017: Applying Principles of Uncertainty within Coastal
9 Hazard Assessments to Better Support Coastal Adaptation. *J. Mar. Sci. Eng.*,
10 doi:10.3390/jmse5030040.
- 11 Stephens, S. A., R. G. Bell, and J. Lawrence, 2018: Developing signals to trigger adaptation to sea-
12 level rise. *Environ. Res. Lett.*, doi:10.1088/1748-9326/aadf96.
- 13 Stern, N., 2007: *The economics of climate change*. Cambridge University Press, Cambridge, UK, and
14 New York, NY, USA,.
- 15 Stern, N., 2013: The Structure of Economic Modeling of the Potential Impacts of Climate Change:
16 Grafting Gross Underestimation of Risk onto Already Narrow Science Models. *J. Econ. Lit.*, **51**,
17 838–859, doi:10.1257/jel.51.3.838.
- 18 Sternberg, T., 2012: Chinese drought, bread and the Arab Spring. *Appl. Geogr.*,
19 doi:10.1016/j.apgeog.2012.02.004.
- 20 —, 2017: Climate hazards in Asian drylands. *Climate Hazard Crises in Asian Societies and*
21 *Environments*.
- 22 Stevanovic, M., and Coauthors, 2016: The impact of high-end climate change on agricultural welfare.
23 *Sci. Adv.*, **2**, e1501452–e1501452, doi:10.1126/sciadv.1501452.
24 <http://advances.sciencemag.org/cgi/doi/10.1126/sciadv.1501452> (Accessed December 9, 2017).
- 25 Stickler, M. M., H. Huntington, A. Haflett, S. Petrova, and I. Bouvier, 2017: Does de facto forest
26 tenure affect forest condition? Community perceptions from Zambia. *For. Policy Econ.*, **85**, 32–
27 45, doi:10.1016/j.forpol.2017.08.014.
- 28 Stoa, R. B., 2015: Droughts, Floods, and Wildfires: Paleo Perspectives on Disaster Law in the
29 Anthropocene. *Georg. Int. Environ. Law Rev.*, **27**, 393–446. [http://commons.cu-](http://commons.cu-portland.edu/lawfaculty)
30 [portland.edu/lawfaculty](http://commons.cu-portland.edu/lawfaculty) (Accessed October 3, 2018).
- 31 Stone, B., J. J. Hess, and H. Frumkin, 2010: Urban form and extreme heat events: are sprawling cities
32 more vulnerable to climate change than compact cities? *Environ. Health Perspect.*, **118**, 1425.
- 33 Stone, J., J. Barclay, P. Simmons, P. D. Cole, S. C. Loughlin, P. Ramón, and P. Mothes, 2014: Risk
34 reduction through community-based monitoring: the vigías of Tungurahua, Ecuador. *J. Appl.*
35 *Volcanol.*, **3**, 11, doi:10.1186/s13617-014-0011-9.
36 <http://appliedvolc.springeropen.com/articles/10.1186/s13617-014-0011-9> (Accessed December
37 11, 2017).
- 38 Storbjörk, S., 2010: “It takes more to get a ship to change course”: Barriers for organizational learning
39 and local climate adaptation in Sweden. *J. Environ. Policy Plan.*,
40 doi:10.1080/1523908X.2010.505414.
- 41 Stringer, L. C., and A. J. Dougill, 2013: Channelling science into policy: Enabling best practices from
42 research on land degradation and sustainable land management in dryland Africa. *J. Environ.*
43 *Manage.*, **114**, 328–335, doi:10.1016/j.jenvman.2012.10.025.
- 44 —, J. C. Dyer, M. S. Reed, A. J. Dougill, C. Twyman, and D. Mkwambisi, 2009: Adaptations to
45 climate change, drought and desertification: local insights to enhance policy in southern Africa.
46 *Environ. Sci. Policy*, **12**, 748–765, doi:10.1016/J.ENVSCI.2009.04.002.
47 <https://www.sciencedirect.com/science/article/pii/S1462901109000604> (Accessed April 7,

- 1 2019).
- 2 Stringer, L. C., and Coauthors, 2018: A New Framework to Enable Equitable Outcomes: Resilience
3 and Nexus Approaches Combined. *Earth's Futur.*, **6**, 902–918, doi:10.1029/2017EF000694.
4 <http://doi.wiley.com/10.1029/2017EF000694> (Accessed April 7, 2019).
- 5 Stupak, I., and K. Raulund-Rasmussen, 2016: Historical, ecological, and governance aspects of
6 intensive forest biomass harvesting in Denmark. *Wiley Interdiscip. Rev. Energy Environ.*, **5**,
7 588–610, doi:10.1002/wene.206. <http://doi.wiley.com/10.1002/wene.206> (Accessed April 12,
8 2019).
- 9 ———, and Coauthors, 2016: A global survey of stakeholder views and experiences for systems needed
10 to effectively and efficiently govern sustainability of bioenergy. *Wiley Interdiscip. Rev. Energy
11 Environ.*, **5**, 89–118, doi:10.1002/wene.166. <http://doi.wiley.com/10.1002/wene.166> (Accessed
12 April 11, 2019).
- 13 Sturm, M., M. . Goldstein, H. P. Huntington, and T. . Douglas, 2017: Using an option pricing
14 approach to evaluate strategic decisions in a rapidly changing climate: Black – Scholes and
15 climate change. *Clim. Change*, **140**.
- 16 Sturrock, R. N., S. J. Frankel, A. V. Brown, P. E. Hennon, J. T. Kliejunas, K. J. Lewis, J. J. Worrall,
17 and A. J. Woods, 2011: Climate change and forest diseases. *Plant Pathol.*, **60**, 133–149,
18 doi:10.1111/j.1365-3059.2010.02406.x.
- 19 Suckall, N., L. C. Stringer, and E. L. Tompkins, 2015: Presenting Triple-Wins? Assessing Projects
20 That Deliver Adaptation, Mitigation and Development Co-benefits in Rural Sub-Saharan Africa.
21 *Ambio*, **44**, 34–41, doi:10.1007/s13280-014-0520-0.
- 22 Sudmeier-Rieux, K., M. Fernández, J. C. Gaillard, L. Guadagno, and M. Jaboyedoff, 2017:
23 Introduction: Exploring Linkages Between Disaster Risk Reduction, Climate Change
24 Adaptation, Migration and Sustainable Development. *Identifying Emerging Issues in Disaster
25 Risk Reduction, Migration, Climate Change and Sustainable Development*.
- 26 Suich, H., C. Howe, and G. Mace, 2015: Ecosystem services and poverty alleviation: a review of the
27 empirical links. *Ecosyst. Serv.*, **12**, 137–147.
- 28 Sumiya, B., 2016: Energy Poverty in Context of Climate Change: What Are the Possible Impacts of
29 Improved Modern Energy Access on Adaptation Capacity of Communities? *Int. J. Environ. Sci.
30 Dev.*, **7**, 7, doi:10.7763/IJESD.2016.V7.744. <http://www.ijesd.org/vol7/744-E0006.pdf>
31 (Accessed April 14, 2019).
- 32 Sun, K., and S. S. Chaturvedi, 2016: Forest conservation and climate change mitigation potential
33 through REDD+ mechanism in Meghalaya, North Eastern India: a review. *Int. J. Sci. Environ.
34 Technol.*, **5**, 3643–3650.
- 35 Sunderlin, W., C. de Sassi, A. Ekaputri, M. Light, and C. Pratama, 2017: REDD+ Contribution to
36 Well-Being and Income Is Marginal: The Perspective of Local Stakeholders. *Forests*, **8**, 125,
37 doi:10.3390/f8040125. <http://www.mdpi.com/1999-4907/8/4/125> (Accessed October 15, 2018).
- 38 Sunderlin, W. D., and Coauthors, 2018: Creating an appropriate tenure foundation for REDD+: The
39 record to date and prospects for the future. *World Dev.*, **106**, 376–392,
40 doi:10.1016/J.WORLDDEV.2018.01.010.
41 <https://www.sciencedirect.com/science/article/pii/S0305750X18300202> (Accessed May 22,
42 2018).
- 43 Sundström, A., 2016: Understanding illegality and corruption in forest governance. *J. Environ.
44 Manage.*, **181**, 779–790, doi:10.1016/j.jenvman.2016.07.020.
45 <http://www.ncbi.nlm.nih.gov/pubmed/27444722> (Accessed November 3, 2018).
- 46 Surminski, S., 2013: Private-sector adaptation to climate risk. *Nat. Clim. Chang.*, **3**, 943–945,
47 doi:10.1038/nclimate2040.
- 48 Surminski, S., 2016: *Submission to the UNFCCC Warsaw International Mechanism by the Loss and*

- 1 *Damage Network*.
- 2 Surminski, S., L. M. Bouwer, and J. Linnerooth-Bayer, 2016: How insurance can support climate
3 resilience. *Nat. Clim. Chang.*, **6**, 333–334, doi:10.1038/nclimate2979.
- 4 Suzuki, R., 2012: *Linking Adaptation and Mitigation through Community Forestry: Case Studies*
5 *from Asia. RECOFTC – The Center for People and Forests*. 80 pp.
- 6 Swanson, A., M. Kosmala, C. Lintott, and C. Packer, 2016: A generalized approach for producing,
7 quantifying, and validating citizen science data from wildlife images. *Conserv. Biol.*, **30**, 520–
8 531, doi:10.1111/cobi.12695. <http://doi.wiley.com/10.1111/cobi.12695> (Accessed October 2,
9 2018).
- 10 Tàbara, J. D., X. Dai, G. Jia, D. McEvoy, H. Neufeldt, A. Serra, S. Werners, and J. J. West, 2010: The
11 climate learning ladder. A pragmatic procedure to support climate adaptation. *Environ. Policy*
12 *Gov.*, **20**, 1–11, doi:10.1002/eet.530.
- 13 Takama, T., S. Tsephel, and F. X. Johnson, 2012: Evaluating the relative strength of product-specific
14 factors in fuel switching and stove choice decisions in Ethiopia. A discrete choice model of
15 household preferences for clean cooking alternatives. *Energy Econ.*, **34**, 1763–1773,
16 doi:10.1016/J.ENERCO.2012.07.001. [https://www.sciencedirect-](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0140988312001375)
17 [com.ezp.sub.su.se/science/article/pii/S0140988312001375](https://www.sciencedirect-com.ezp.sub.su.se/science/article/pii/S0140988312001375) (Accessed May 23, 2018).
- 18 Tallis, H., P. Kareiva, M. Marvier, and A. Chang, 2008: An ecosystem services framework to support
19 both practical conservation and economic development. *Proc. Natl. Acad. Sci.*, **105**, 9457–9464.
- 20 Tanner, T., and Coauthors, 2015: Livelihood resilience in the face of climate change. *Nat. Clim.*
21 *Chang.*, doi:10.1038/nclimate2431.
- 22 Tao, H.-H., J. L. Snaddon, E. M. Slade, J.-P. Caliman, R. H. Widodo, Suhardi, and K. J. Willis, 2017:
23 Long-term crop residue application maintains oil palm yield and temporal stability of
24 production. *Agron. Sustain. Dev.*, **37**, 33, doi:10.1007/s13593-017-0439-5.
25 <http://link.springer.com/10.1007/s13593-017-0439-5> (Accessed April 15, 2019).
- 26 Taylor, R. G., and Coauthors, 2013: Ground water and climate change. *Nat. Clim. Chang.*, **3**, 322–
27 329, doi:10.1038/nclimate1744. <http://www.nature.com/articles/nclimate1744> (Accessed May
28 18, 2018).
- 29 TEEB, T., 2009: *The Economics of Ecosystems and Biodiversity for National and International*
30 *Policy Makers—Summary: Responding to the Value of Nature 2009*.
- 31 Temper, L., and J. Martinez-Alier, 2013: The god of the mountain and Godavarman: Net Present
32 Value, indigenous territorial rights and sacredness in a bauxite mining conflict in India. *Ecol.*
33 *Econ.*, **96**, 79–87.
- 34 Tengberg, A., and S. Valencia, 2018: Integrated approaches to natural resources management-Theory
35 and practice. *L. Degrad. Dev.*, **29**, 1845–1857, doi:10.1002/ldr.2946.
36 <http://doi.wiley.com/10.1002/ldr.2946> (Accessed February 25, 2019).
- 37 —, F. Radstake, K. Zhang, and B. Dunn, 2016: Scaling up of Sustainable Land Management in the
38 Western People’s Republic of China: Evaluation of a 10-Year Partnership. *L. Degrad. Dev.*, **27**,
39 134–144, doi:10.1002/ldr.2270.
- 40 Tengö, M., E. S. Brondizio, T. Elmqvist, P. Malmer, and M. Spierenburg, 2014: Connecting diverse
41 knowledge systems for enhanced ecosystem governance: The multiple evidence base approach.
42 *Ambio*, **43**, 579–591, doi:10.1007/s13280-014-0501-3.
- 43 Tennigkeit, T., and W. Andreas, 2008: *Working Paper An Assessment of the Potential for Carbon*
44 *Finance* *in* *Rangelands*.
45 <http://www.worldagroforestry.org/downloads/Publications/PDFS/WP15892.pdf> (Accessed
46 November 3, 2018).
- 47 Termeer, C. J. A. M., A. Dewulf, and G. R. Biesbroek, 2017: Transformational change: governance

- 1 interventions for climate change adaptation from a continuous change perspective. *J. Environ.*
2 *Plan. Manag.*, doi:10.1080/09640568.2016.1168288.
- 3 Terrazas, W. C. M., V. de S. Sampaio, D. B. de Castro, R. C. Pinto, B. C. de Albuquerque, M.
4 Sadahiro, R. A. dos Passos, and J. U. Braga, 2015: Deforestation, drainage network, indigenous
5 status, and geographical differences of malaria in the State of Amazonas. *Malar. J.*, **14**, 379,
6 doi:10.1186/s12936-015-0859-0. <http://www.malariajournal.com/content/14/1/379> (Accessed
7 December 29, 2017).
- 8 Tessler, Z. D., C. J. Vörösmarty, M. Grossberg, I. Gladkova, H. Aizenman, J. P. M. Syvitski, and E.
9 Foufoula-Georgiou, 2015: Profiling risk and sustainability in coastal deltas of the world. *Science*
10 (80-), doi:10.1126/science.aab3574.
- 11 Thaker, M, Zambre, A. Bhosale, H., 2018: Wind farms have cascading impacts on ecosystems across
12 trophic levels. *Nat. Ecol. Evol.*,.
- 13 Thaler, R. H., and C. R. Sunstein, 2008: *Nudge: Improving decisions about health, wealth, and*
14 *happiness.* 1-293 pp. [https://www.scopus.com/inward/record.uri?eid=2-s2.0-](https://www.scopus.com/inward/record.uri?eid=2-s2.0-84903035283&partnerID=40&md5=dcaa45ffc4e37961daa79f6dadadfb71)
15 [84903035283&partnerID=40&md5=dcaa45ffc4e37961daa79f6dadadfb71](https://www.scopus.com/inward/record.uri?eid=2-s2.0-84903035283&partnerID=40&md5=dcaa45ffc4e37961daa79f6dadadfb71).
- 16 Thamo, T., and D. J. Pannell, 2016: Challenges in developing effective policy for soil carbon
17 sequestration: perspectives on additionality, leakage, and permanence. *Clim. Policy*, **16**, 973–
18 992, doi:10.1080/14693062.2015.1075372.
19 <https://www.tandfonline.com/doi/full/10.1080/14693062.2015.1075372> (Accessed May 22,
20 2018).
- 21 Theisen, O. M., H. Holtermann, and H. Buhaug, 2011: Climate Wars? Assessing the Claim That
22 Drought Breeds Conflict. *Int. Secur.*, doi:10.1162/isec_a_00065.
- 23 Theriault, V., M. Smale, and H. Haider, 2017a: How Does Gender Affect Sustainable Intensification
24 of Cereal Production in the West African Sahel? Evidence from Burkina Faso. *World Dev.*, **92**,
25 177–191, doi:10.1016/J.WORLDDEV.2016.12.003.
26 <https://www.sciencedirect.com/science/article/pii/S0305750X16305575> (Accessed April 2,
27 2019).
- 28 —, —, and —, 2017b: How Does Gender Affect Sustainable Intensification of Cereal
29 Production in the West African Sahel? Evidence from Burkina Faso. *World Dev.*, **92**, 177–191,
30 doi:10.1016/J.WORLDDEV.2016.12.003.
- 31 Thomas, A., and L. Benjamin, 2017: Policies and mechanisms to address climate-induced migration
32 and displacement in Pacific and Caribbean small island developing states. *Int. J. Clim. Chang.*
33 *Strateg. Manag.*, doi:10.1108/IJCCSM-03-2017-0055.
- 34 Thomas, D. H. L., 1996: Fisheries tenure in an African floodplain village and the implications for
35 management. *Hum. Ecol.*, **24**, 287–313.
- 36 Thompson-Hall, M., E. R. Carr, and U. Pascual, 2016: Enhancing and expanding intersectional
37 research for climate change adaptation in agrarian settings. *Ambio*, **45**, 373–382,
38 doi:10.1007/s13280-016-0827-0. <http://link.springer.com/10.1007/s13280-016-0827-0>
39 (Accessed April 2, 2019).
- 40 Thompson, I., B. Mackey, S. McNulty, and A. Mosseler, 2009: Forest resilience, biodiversity, and
41 climate change. *Secretariat of the Convention on Biological Diversity, Montreal. Technical*
42 *Series no. 43. 1-67.*, Vol. 43 of, 1–67.
- 43 Thorén, H., and L. Olsson, 2017: Is resilience a normative concept? *Resilience*,
44 doi:10.1080/21693293.2017.1406842.
- 45 Tidwell, J. H., and G. L. Allan, 2001: Fish as food: aquaculture’s contribution: Ecological and
46 economic impacts and contributions of fish farming and capture fisheries. *EMBO Rep.*, **2**, 958–
47 963.
- 48 Tierney, J. E., C. C. Ummenhofer, and P. B. DeMenocal, 2015: Past and future rainfall in the Horn of

- 1 Africa. *Sci. Adv.*, doi:10.1126/sciadv.1500682.
- 2 Tigchelaar, M., D. Battisti, R. . Naylor, and D. . Ray, 2018: Probability of globally synchronized
3 maize production shocks. *Proc. Natl. Acad. Sci.*, **115**, 6644–6649.
- 4 Timberlake, T. J., and C. A. Schultz, 2017: Policy, practice, and partnerships for climate change
5 adaptation on US national forests. *Clim. Change*, doi:10.1007/s10584-017-2031-z.
- 6 Tittonell, P., 2014: Livelihood strategies, resilience and transformability in African agroecosystems.
7 *Agric. Syst.*, **126**, 3–14, doi:10.1016/j.agry.2013.10.010.
- 8 Tjaden, N. B., J. E. Suk, D. Fischer, S. M. Thomas, C. Beierkuhnlein, and J. C. Semenza, 2017:
9 Modelling the effects of global climate change on Chikungunya transmission in the 21 st
10 century. *Sci. Rep.*, doi:10.1038/s41598-017-03566-3.
- 11 Tompkins, E. L., and W. N. Adger, 2004: Does Adaptive Management of Natural Resources Enhance
12 Resilience to Climate Change? *Ecol. Soc.*, **9**, 10.
- 13 Torvanger, A., 2019a: Governance of bioenergy with carbon capture and storage (BECCS):
14 accounting, rewarding, and the Paris agreement. *Clim. Policy*, **19**, 329–341,
15 doi:10.1080/14693062.2018.1509044.
16 <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1509044> (Accessed April 11,
17 2019).
- 18 —, 2019b: Governance of bioenergy with carbon capture and storage (BECCS): accounting,
19 rewarding, and the Paris agreement. *Clim. Policy*, **19**, 329–341,
20 doi:10.1080/14693062.2018.1509044.
- 21 Tóth, G., T. Hermann, M. R. da Silva, and L. Montanarella, 2018: Monitoring soil for sustainable
22 development and land degradation neutrality. *Environ. Monit. Assess.*, doi:10.1007/s10661-017-
23 6415-3.
- 24 Totin, E., A. Segnon, M. Schut, H. Affognon, R. Zougmore, T. Rosenstock, and P. Thornton, 2018:
25 Institutional Perspectives of Climate-Smart Agriculture: A Systematic Literature Review.
26 *Sustainability*, **10**, 1990, doi:10.3390/su10061990. <http://www.mdpi.com/2071-1050/10/6/1990>
27 (Accessed October 15, 2018).
- 28 Toulmin, C., and J. Quan, 2000: *Evolving land rights, policy and tenure in Africa*. IIED, 324 pp.
29 www.iied.org (Accessed May 23, 2018).
- 30 Travis, W. R., 2013: Design of a severe climate change early warning system. *Weather Clim. Extrem.*,
31 doi:10.1016/j.wace.2013.10.006.
- 32 Tribbia, J., and S. C. Moser, 2008: More than information: what coastal managers need to plan for
33 climate change. *Environ. Sci. Policy*, **11**, 315–328, doi:10.1016/J.ENVSCI.2008.01.003.
34 <https://www.sciencedirect.com/science/article/pii/S1462901108000130> (Accessed May 24,
35 2018).
- 36 Trieb, F., H. Müller-Steinhagen, and J. Kern, 2011: Financing concentrating solar power in the
37 Middle East and North Africa—Subsidy or investment? *Energy Policy*, **39**, 307–317.
- 38 Tschakert, P., 2007: Views from the vulnerable: Understanding climatic and other stressors in the
39 Sahel. *Glob. Environ. Chang.*, **17**, 381–396.
- 40 Tucker Lima, J. M., A. Vittor, S. Rifai, and D. Valle, 2017: Does deforestation promote or inhibit
41 malaria transmission in the Amazon? A systematic literature review and critical appraisal of
42 current evidence. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, **372**, 20160125,
43 doi:10.1098/rstb.2016.0125. <http://www.ncbi.nlm.nih.gov/pubmed/28438914> (Accessed
44 December 29, 2017).
- 45 Tularam, G., and M. Krishna, 2009: Long term consequences of groundwater pumping in Australia: A
46 review of impacts around the globe. *J. Appl. Sci. Environ. Sanit.*, **4**, 151–166. [https://research-
47 repository.griffith.edu.au/handle/10072/29294](https://research-repository.griffith.edu.au/handle/10072/29294) (Accessed March 29, 2019).

- 1 Turkelboom, F., and Coauthors, 2018: When we cannot have it all: Ecosystem services trade-offs in
2 the context of spatial planning. *Ecosyst. Serv.*, **29**, 566–578.
- 3 Turnhout, E., K. Neves, and E. de Lijster, 2014: ‘Measurementality’ in biodiversity governance:
4 knowledge, transparency, and the Intergovernmental Science-Policy Platform on Biodiversity
5 and Ecosystem Services (IPBES). *Environ. Plan. A*, **46**, 581–597.
- 6 von Uexkull, N., M. Croicu, H. Fjelde, and H. Buhaug, 2016: Civil conflict sensitivity to growing-
7 season drought. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1607542113.
- 8 Ulrich-Schad, J.D., Garcia de Jalon, S., Babin, N., Paper, A., P. L. S., 2017: Measuring and
9 understanding agricultural producers’ adoption of nutrient best management practices. *J. Soil
10 Water Conserv.*, **72**, 506–518, doi:10.2489/jswc.72.5.506.
- 11 Umukoro, N., 2013: Poverty and Social Protection in Nigeria. *J. Dev. Soc.*, **29**, 305–322,
12 doi:10.1177/0169796X13494281.
- 13 UNCCD/Science-Policy-Interface, 2016: Land in Balance The Scientific Conceptual Framework for
14 Land Degradation Neutrality The principles of LDN. *Sci. Br.*,
- 15 UNCCD, 2015: *Land Degradation Neutrality: The Target Setting Programme*. 22 pp.
- 16 UNDP, Governance for Sustainable Human Development, New York: UNDP, 1997, pp. 2-3.
- 17 UNEP, 2009: *Secretariat of the Convention on Biological Diversity STATEMENT by AHMED
18 DJOGHLAF EXECUTIVE SECRETARY at the Meeting of Steering Committee Global Form on
19 Oceans, Coasts and Islands*. <http://www.cbd.int/secretariat@cbd.int> (Accessed October 31,
20 2018).
- 21 —, 2016: *The Adaptation Finance Gap Report 2016*.
- 22 UNFCCC, 2007: Climate Change: Impacts, Vulnerabilities and Adaptation in Developing Countries.
23 *United Nations Framew. Conv. Clim. Chang.*, 68, doi:10.1029/2005JD006289.
- 24 —, 2018a: *Paris Rulebook: Proposal by the President, Informal compilation of L-documents*.
25 UNFCCC, Katowice, [https://unfccc.int/sites/default/files/resource/Informal
26 Compilation_proposal by the President_rev.pdf](https://unfccc.int/sites/default/files/resource/Informal_Compilation_proposal_by_the_President_rev.pdf) (Accessed March 17, 2019).
- 27 —, 2018b: *Paris Rulebook: Proposal by the President, Informal compilation of L-documents*.
28 UNFCCC, Katowice,.
- 29 UNFCCC (United Nations Framework Convention on Climate Change), 2016: Paris Agreement.
30 *Paris Agreem. - Pre 2020 Action*,.
- 31 United Nations Environment Programme, 2017: *The Emissions Gap Report 2017*. 1-86 pp.
- 32 Urwin, K., and A. Jordan, 2008: Does public policy support or undermine climate change adaptation?
33 Exploring policy interplay across different scales of governance. *Glob. Environ. Chang.*, **18**,
34 180–191, doi:10.1016/j.gloenvcha.2007.08.002.
- 35 Usher, P. J., 2000: *Traditional Ecological Knowledge in Environmental Assessment and Management*.
36 183-193 pp.
37 <http://citeseerx.ist.psu.edu/viewdoc/download?doi=10.1.1.474.5055&rep=rep1&type=pdf>
38 (Accessed April 15, 2019).
- 39 Uzun, B., and M. Cete, 2004: A Model for Solving Informal Settlement Issues in Developing
40 Countries. *Planning, Valuat. Environ.*,.
- 41 Valatin, G., D. Moseley, and N. Dandy, 2016: Insights from behavioural economics for forest
42 economics and environmental policy: Potential nudges to encourage woodland creation for
43 climate change mitigation and adaptation? *For. Policy Econ.*, **72**, 27–36,
44 doi:10.1016/j.forpol.2016.06.012.
- 45 Valérie Masson-Delmotte, H.-O. Pörtner, J. Skea, P. Zhai, D. Roberts, and P. R. Shukla, 2018:
46 *Intergovernmental Panel on Climate Change Global Warming of 1.5o C Report Summary for*

- 1 *Policy Makers.*
- 2 Valipour, A., T. Plieninger, Z. Shakeri, H. Ghazanfari, M. Namiranian, and M. J. Lexer, 2014:
3 Traditional silvopastoral management and its effects on forest stand structure in northern Zagros,
4 Iran. *For. Ecol. Manage.*, **327**, 221–230, doi:10.1016/j.foreco.2014.05.004.
- 5 Vandersypen, K., A. C. T. Keita, Y. Coulibaly, D. Raes, and J. Y. Jamin, 2007: Formal and informal
6 decision making on water management at the village level: A case study from the Office du
7 Niger irrigation scheme (Mali). *Water Resour. Res.*, doi:10.1029/2006WR005132.
- 8 Vanmaercke, M., and Coauthors, 2016a: How fast do gully headcuts retreat? *Earth-Science Rev.*, **154**,
9 336–355, doi:10.1016/J.EARSCIREV.2016.01.009.
10 <https://www.sciencedirect.com/science/article/pii/S0012825216300083> (Accessed April 14,
11 2019).
- 12 —, and Coauthors, 2016b: How fast do gully headcuts retreat? *Earth-Science Rev.*, **154**, 336–355,
13 doi:10.1016/J.EARSCIREV.2016.01.009.
- 14 Velthof, G. L., J. P. Lesschen, J. Webb, S. Pietrzak, Z. Miatkowski, M. Pinto, J. Kros, and O.
15 Oenema, 2014: The impact of the Nitrates Directive on nitrogen emissions from agriculture in
16 the EU-27 during 2000-2008. *Sci. Total Environ.*, doi:10.1016/j.scitotenv.2013.04.058.
- 17 Vent, O., Sabarmatee, and N. Uphoff, 2017: The System of Rice Intensification and its impacts on
18 women: Reducing pain, discomfort, and labor in rice farming while enhancing households' food
19 security. *Women in Agriculture Worldwide: Key issues and practical approaches*, A. Fletcher
20 and W. Kubik, Eds., Routledge, London and New York, 55–76
21 [https://books.google.co.za/books?id=myeTDAAAQBAJ&pg=PT7&lpg=PT7&dq=The+System
22 +of+Rice+Intensification+and+its+impacts+on+women:+Reducing+pain,+discomfort,+and+lab
23 or+in+rice+farming+while+enhancing+households'+food+security.&source=bl&ots=43HYQDJ
24 eR](https://books.google.co.za/books?id=myeTDAAAQBAJ&pg=PT7&lpg=PT7&dq=The+System+of+Rice+Intensification+and+its+impacts+on+women:+Reducing+pain,+discomfort,+and+lab+or+in+rice+farming+while+enhancing+households'+food+security.&source=bl&ots=43HYQDJ) (Accessed April 4, 2019).
- 25 Venton, C. C., 2018: *The economics of resilience to drought.*
26 [https://www.usaid.gov/sites/default/files/documents/1867/Kenya_Economics_of_Resilience_Fin
27 al_Jan_4_2018_-_BRANDED.pdf](https://www.usaid.gov/sites/default/files/documents/1867/Kenya_Economics_of_Resilience_Final_Jan_4_2018_-_BRANDED.pdf).
- 28 Venton, C. C. C., C. Fitzgibbon, T. Shitarek, L. Coulter, and O. Dooley, 2012: *The Economics of*
29 *Early Response and Disaster Resilience: Lessons from Kenya and Ethiopia.* 1-84 pp.
30 [http://collection.europarchive.org/tna/20121003151823/http://dfid.gov.uk/Documents/publicatio
31 ns1/Econ-Ear-Rec-Res-Full-Report.pdf](http://collection.europarchive.org/tna/20121003151823/http://dfid.gov.uk/Documents/publications1/Econ-Ear-Rec-Res-Full-Report.pdf).
- 32 Verburg, P. H., and Coauthors, 2015: Land system science and sustainable development of the earth
33 system: A global land project perspective. *Anthropocene*, **12**, 29–41,
34 doi:10.1016/j.ancene.2015.09.004.
- 35 Verbyla, D., 2011: Browning boreal forests of western North America. *Environ. Res. Lett.*, **6**, 41003.
- 36 Verchot, L. V., and Coauthors, 2007: Climate change: linking adaptation and mitigation through
37 agroforestry. *Mitig. Adapt. Strateg. Glob. Chang.*, **12**, 901–918.
- 38 Verdegem, M. C. J., and R. H. Bosma, 2009: Water withdrawal for brackish and inland aquaculture,
39 and options to produce more fish in ponds with present water use. *Water Policy*, **11**, 52–68.
- 40 Vergara-Asenjo, G., and C. Potvin, 2014: Forest protection and tenure status: The key role of
41 indigenous peoples and protected areas in Panama. *Glob. Environ. Chang.*, **28**, 205–215,
42 doi:10.1016/J.GLOENVCHA.2014.07.002.
43 <https://www.sciencedirect.com/science/article/pii/S0959378014001289> (Accessed April 9,
44 2019).
- 45 Verma, S., D. A. Kampman, P. van der Zaag, and A. Y. Hoekstra, 2009: Going against the flow: a
46 critical analysis of inter-state virtual water trade in the context of India's National River Linking
47 Program. *Phys. Chem. Earth, Parts A/B/C*, **34**, 261–269.
- 48 Verschuuren, J., 2017: Towards a Regulatory Design for Reducing Emissions from Agriculture:

- 1 Lessons from Australia's Carbon Farming Initiative. *Clim. Law*, **7**, 1–51, doi:10.1163/18786561-
2 00701001. <http://booksandjournals.brillonline.com/content/journals/10.1163/18786561->
3 00701001 (Accessed October 8, 2018).
- 4 Vervoort, J., and A. Gupta, 2018: Anticipating climate futures in a 1.5 °C era: the link between
5 foresight and governance. *Curr. Opin. Environ. Sustain.*, doi:10.1016/j.cosust.2018.01.004.
- 6 Verweij, M., and Coauthors, 2006: Clumsy solutions for a complex world: The case of climate
7 change. *Public Adm.*, **84**, 817–843, doi:10.1111/j.1467-9299.2006.00614.x.
- 8 Vijge, M. J., and A. Gupta, 2014: Framing REDD+ in India: Carbonizing and centralizing Indian
9 forest governance? *Environ. Sci. Policy*, **38**, 17–27.
- 10 Villagra, P., and C. Quintana, 2017: Disaster Governance for Community Resilience in Coastal
11 Towns: Chilean Case Studies. *Int. J. Environ. Res. Public Health*, **14**, 1063,
12 doi:10.3390/ijerph14091063. <http://www.mdpi.com/1660-4601/14/9/1063>.
- 13 Vincent, K., S. Besson, T. Cull, and C. Menzel, 2018: Sovereign insurance to incentivize the shift
14 from disaster response to adaptation to climate change—African Risk Capacity's Extreme
15 Climate Facility. *Clim. Dev.*, **10**, 385–388, doi:10.1080/17565529.2018.1442791.
- 16 Vira, B., B. Adams, C. Agarwal, S. Badiger, R. a Hope, J. Krishnaswamy, and C. Kumar, 2012:
17 Negotiating trade-offs. *Econ. Polit. Wkly.*, **47**, 67.
- 18 Vörösmarty, C. J., and Coauthors, 2010: Global threats to human water security and river
19 biodiversity. *Nature*, **467**, 555–561, doi:10.1038/nature09440.
20 <http://www.ncbi.nlm.nih.gov/pubmed/20882010> (Accessed May 25, 2018).
- 21 Voß, J.-P., and N. Amelung, 2016: Innovating public participation methods: Technoscientization and
22 reflexive engagement. *Soc. Stud. Sci.*, **46**, 749–772, doi:10.1177/0306312716641350.
- 23 Voß, J. P., and A. Simons, 2018: A novel understanding of experimentation in governance: co-
24 producing innovations between “lab” and “field.” *Policy Sci.*, doi:10.1007/s11077-018-9313-9.
- 25 Waas, T., J. Hugé, T. Block, T. Wright, F. Benitez-Capistros, and A. Verbruggen, 2014: Sustainability
26 assessment and indicators: Tools in a decision-making strategy for sustainable development.
27 *Sustain.*, **6**, 5512–5534, doi:10.3390/su6095512.
- 28 Wada, Y., L. P. H. van Beek, C. M. van Kempen, J. W. T. M. Reckman, S. Vasak, and M. F. P.
29 Bierkens, 2010: Global depletion of groundwater resources. *Geophys. Res. Lett.*, **37**, n/a-n/a,
30 doi:10.1029/2010GL044571. <http://doi.wiley.com/10.1029/2010GL044571> (Accessed May 18,
31 2018).
- 32 Wada, Y., A. K. Gain, and C. Giupponi, 2016: Measuring global water security towards sustainable
33 development goals. *Environ. Res. Lett.*, **11**, 2–13.
- 34 Waddock, S., 2013: The wicked problems of global sustainability need wicked (good) leaders and
35 wicked (good) collaboraative solutions. *J. Manag. Glob. Sustain.*, **1**, 91–111,
36 doi:10.13185/JM2013.01106.
- 37 Wagenbrenner, N. S., M. J. Germino, B. K. Lamb, P. R. Robichaud, and R. B. Foltz, 2013: Wind
38 erosion from a sagebrush steppe burned by wildfire: Measurements of PM10 and total horizontal
39 sediment flux. *Aeolian Res.*, **10**, 25–36, doi:10.1016/j.aeolia.2012.10.003.
40 <http://dx.doi.org/10.1016/j.aeolia.2012.10.003>.
- 41 Wagner, G., 2013: Carbon Cap and Trade. *Encycl. Energy, Nat. Resour. Environ. Econ.*, **1–3**, 1–5,
42 doi:10.1016/B978-0-12-375067-9.00071-1.
- 43 Waite, S. H., 2011: Blood Forests: Post Lacey Act, Why Cohesive Global Governance Is Essential to
44 Extinguish the Market for Illegally Harvested Timber. *Seattle J. Environ. Law*, **2**.
45 <https://heinonline.org/HOL/Page?handle=hein.journals/sjel4&id=314&div=9&collection=journals>
46 [ls](https://heinonline.org/HOL/Page?handle=hein.journals/sjel4&id=314&div=9&collection=journals) (Accessed November 2, 2018).
- 47 Walker, W., and Coauthors, 2014: Forest carbon in Amazonia: the unrecognized contribution of

- 1 indigenous territories and protected natural areas. *Carbon Manag.*, **5**, 479–485,
2 doi:10.1080/17583004.2014.990680.
3 <http://www.tandfonline.com/doi/full/10.1080/17583004.2014.990680> (Accessed April 11,
4 2019).
- 5 Walter, A., J. E. A. Seabra, P. G. Machado, B. de Barros Correia, and C. O. F. de Oliveira, 2018:
6 Sustainability of Biomass. *Biomass and Green Chemistry*, Springer International Publishing,
7 Cham, 191–219 http://link.springer.com/10.1007/978-3-319-66736-2_8 (Accessed April 8,
8 2019).
- 9 Wam, H. K., N. Bunnefeld, N. Clarke, and O. Hofstad, 2016: Conflicting interests of ecosystem
10 services: Multi-criteria modelling and indirect evaluation of trade-offs between monetary and
11 non-monetary measures. *Ecosyst. Serv.*, doi:10.1016/j.ecoser.2016.10.003.
- 12 Wang, C., J. Liu, J. Shen, D. Chen, Y. Li, B. Jiang, and J. Wu, 2018: Effects of biochar amendment
13 on net greenhouse gas emissions and soil fertility in a double rice cropping system: A 4-year
14 field experiment. *Agric. Ecosyst. Environ.*, **262**, 83–96, doi:10.1016/J.AGEE.2018.04.017.
15 [https://www-sciencedirect-
16 com.ezp.sub.su.se/science/article/pii/S016788091830166X?via%3Dihub](https://www-sciencedirect-com.ezp.sub.su.se/science/article/pii/S016788091830166X?via%3Dihub) (Accessed April 15,
17 2019).
- 18 Wang, S., and B. Fu, 2013: Trade-offs between forest ecosystem services. *For. Policy Econ.*,
19 doi:10.1016/j.forpol.2012.07.014.
- 20 Wang, X., G. Zhou, D. Zhong, X. Wang, Y. Wang, Z. Yang, L. Cui, and G. Yan, 2016: Life-table
21 studies revealed significant effects of deforestation on the development and survivorship of
22 *Anopheles minimus* larvae. *Parasit. Vectors*, **9**, 323, doi:10.1186/s13071-016-1611-5.
23 <http://parasitesandvectors.biomedcentral.com/articles/10.1186/s13071-016-1611-5> (Accessed
24 December 29, 2017).
- 25 Ward, F. A., and M. Pulido-Velazquez, 2008: Water conservation in irrigation can increase water use.
26 *Proc. Natl. Acad. Sci.*, **105**, 18215–18220.
- 27 Ward, P. S., 2016: Transient poverty, poverty dynamics, and vulnerability to poverty: An empirical
28 analysis using a balanced panel from rural China. *World Dev.*, **78**, 541–553,
29 doi:10.1016/j.worlddev.2015.10.022.
- 30 Warner, K., 2018: Coordinated approaches to large-scale movements of people: contributions of the
31 Paris Agreement and the Global Compacts for migration and on refugees. *Popul. Environ.*, **39**,
32 384–401, doi:<https://doi.org/10.1007/s11111-018-0299-1>.
- 33 Warner, K., and T. Afifi, 2011: Special issue ENVIRONMENTALLY INDUCED MIGRATION IN
34 THE CONTEXT OF SOCIAL VULNERABILITY. *Int. Migr.*, doi:10.1111/j.1468-
35 2435.2011.00697.x.
- 36 Warner, K., and T. Afifi, 2014: Where the rain falls: Evidence from 8 countries on how vulnerable
37 households use migration to manage the risk of rainfall variability and food insecurity. *Clim.
38 Dev.*, doi:10.1080/17565529.2013.835707.
- 39 —, K. Van Der Geest, S. Kreft, S. Huq, S. Harmeling, K. Kusters, and A. De Sherbinin, 2012:
40 *Evidence from the frontlines of climate change: Loss and damage to communities despite coping
41 and adaptation.*
- 42 —, Z. Zommers, A. Wreford, M. Hurlbert, D. Viner, J. Scantlan, K. Halsey, and C. Tamang, 2018:
43 Characteristics of transformational adaptation in land-society-climate interactions.
44 *Sustainability*,.
- 45 —, and Coauthors, 2019: Characteristics of Transformational Adaptation in Climate-Land-Society
46 Interactions. *Sustainability*, doi:10.3390/su11020356.
- 47 Wathore, R., K. Mortimer, and A. P. Grieshop, 2017: In-Use Emissions and Estimated Impacts of
48 Traditional, Natural- and Forced-Draft Cookstoves in Rural Malawi. *Environ. Sci. Technol.*, **51**,

- 1 1929–1938, doi:10.1021/acs.est.6b05557. <http://pubs.acs.org/doi/10.1021/acs.est.6b05557>
2 (Accessed April 15, 2019).
- 3 Watts, N., and Coauthors, 2015: Health and climate change: policy responses to protect public health.
4 *Lancet*, **386**, 1861–1914, doi:10.1016/S0140-6736(15)60854-6.
5 <http://linkinghub.elsevier.com/retrieve/pii/S0140673615608546> (Accessed December 29, 2017).
- 6 —, and Coauthors, 2018: The 2018 report of the Lancet Countdown on health and climate change:
7 shaping the health of nations for centuries to come. *Lancet*, doi:10.1016/S0140-6736(18)32594-
8 7.
- 9 Weichselgartner, J., and I. Kelman, 2015: Geographies of resilience: Challenges and opportunities of
10 a descriptive concept. *Prog. Hum. Geogr.*, doi:10.1177/0309132513518834.
- 11 Weick, K. E., and K. M. Sutcliffe, 2001: *Managing the Unexpected. Resilient performance in a time*
12 *of change*. 200 pp.
- 13 Weitz, N., H. Carlsen, M. Nilsson, and K. Skånberg, 2017a: Towards systemic and contextual priority
14 setting for implementing the 2030 Agenda. *Sustainability Science*.
- 15 —, C. Strambo, E. Kemp-Benedict, and M. Nilsson, 2017b: Closing the governance gaps in the
16 water-energy-food nexus: Insights from integrative governance. *Glob. Environ. Chang.*,
17 doi:10.1016/j.gloenvcha.2017.06.006.
- 18 Welcomme, R. L., I. G. Cowx, D. Coates, C. Béné, S. Funge-Smith, A. Halls, and K. Lorenzen, 2010:
19 Inland capture fisheries. *Philos. Trans. R. Soc. London B Biol. Sci.*, **365**, 2881–2896.
- 20 Wellesley, L., F. Preston, J. Lehne, and R. Bailey, 2017: Chokepoints in global food trade: Assessing
21 the risk. *Res. Transp. Bus. Manag.*, doi:10.1016/j.rtbm.2017.07.007.
- 22 Wenkel, K.-O., M. Berg, W. Mirschel, R. Wieland, C. Nendel, and B. Köstner, 2013: LandCaRe DSS
23 – An interactive decision support system for climate change impact assessment and the analysis
24 of potential agricultural land use adaptation strategies. *J. Environ. Manage.*, **127**, S168–S183,
25 doi:10.1016/J.JENVMAN.2013.02.051.
26 <http://www.sciencedirect.com.ezp.sub.su.se/science/article/pii/S0301479713001497> (Accessed
27 December 22, 2017).
- 28 West, T. A. P., 2016: Indigenous community benefits from a de-centralized approach to REDD+ in
29 Brazil. *Clim. Policy*, **16**, 924–939, doi:10.1080/14693062.2015.1058238.
30 <http://www.tandfonline.com/doi/full/10.1080/14693062.2015.1058238> (Accessed October 15,
31 2018).
- 32 Westberg, C. J., and F. X. Johnson, 2013: *The Path Not Yet Taken: Bilateral Agreements to Promote*
33 *Sustainable Biofuels under the EU Renewable Energy Directive Stockholm Environment*
34 *Institute, Working Paper 2013-02*. Stockholm, www.sei.org (Accessed April 13, 2019).
- 35 Westerling, A. L., H. G. Hidalgo, D. R. Cayan, and T. W. Swetnam, 2006: Warming and Earlier
36 Spring Increase Western U.S. Forest Wildfire Activity. *Science (80-.)*, **313**, 940–943,
37 doi:10.1126/SCIENCE.1128834.
- 38 Weyant, C. L., and Coauthors, 2019a: Emission Measurements from Traditional Biomass Cookstoves
39 in South Asia and Tibet. *Environ. Sci. Technol.*, **53**, 3306–3314, doi:10.1021/acs.est.8b05199.
40 <http://pubs.acs.org/doi/10.1021/acs.est.8b05199> (Accessed April 14, 2019).
- 41 —, and Coauthors, 2019b: Emission Measurements from Traditional Biomass Cookstoves in South
42 Asia and Tibet. *Environ. Sci. Technol.*, **53**, 3306–3314, doi:10.1021/acs.est.8b05199.
- 43 Wheaton, E., and S. Kulshreshtha, 2017: Environmental sustainability of agriculture stressed by
44 changing extremes of drought and excess moisture: A conceptual review. *Sustain.*, **9**,
45 doi:10.3390/su9060970.
- 46 Wheeler, T., and J. Von Braun, 2013: Climate change impacts on global food security. *Science (80-.*
47 *)*, doi:10.1126/science.1239402.

- 1 Whitaker, J., and Coauthors, 2018: Consensus, uncertainties and challenges for perennial bioenergy
2 crops and land use. *GCB Bioenergy*, doi:10.1111/gcbb.12488.
- 3 White, B., S. M. Borrás, R. Hall, I. Scoones, and W. Wolford, 2012: The new enclosures: Critical
4 perspectives on corporate land deals. *J. Peasant Stud.*, **39**, 619–647,
5 doi:10.1080/03066150.2012.691879.
- 6 White, J., and J. Morton, 2005: Mitigating impacts of HIV/AIDS on rural livelihoods: NGO
7 experiences in sub-Saharan Africa. *Dev. Pract.*, **15**, 186–199, doi:10.1080/09614520500041757.
8 <http://www.tandfonline.com/doi/abs/10.1080/09614520500041757> (Accessed April 7, 2019).
- 9 Whitmee, S., and Coauthors, 2015: Safeguarding human health in the Anthropocene epoch: Report of
10 the Rockefeller Foundation-Lancet Commission on planetary health. *Lancet*,
11 doi:10.1016/S0140-6736(15)60901-1.
- 12 Wiebe, K., and Coauthors, 2015a: Climate change impacts on agriculture in 2050 under a range of
13 plausible socioeconomic and emissions scenarios. *Environ. Res. Lett.*, **10**, 085010,
14 doi:10.1088/1748-9326/10/8/085010. [http://stacks.iop.org/1748-
15 9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b](http://stacks.iop.org/1748-9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b) (Accessed April 12,
16 2019).
- 17 —, and Coauthors, 2015b: Climate change impacts on agriculture in 2050 under a range of
18 plausible socioeconomic and emissions scenarios. *Environ. Res. Lett.*, **10**, 085010,
19 doi:10.1088/1748-9326/10/8/085010.
- 20 Wiebe, K., and Coauthors, 2018: *Scenario development and foresight analysis: Exploring options to
21 inform choices*.
- 22 Wiggering, H., and U. Steinhardt, 2015: A conceptual model for site-specific agricultural land-use.
23 *Ecol. Modell.*, **295**, 42–46, doi:10.1016/j.ecolmodel.2014.08.011.
- 24 Wilby, R. L., and S. Dessai, 2010: Robust adaptation to climate change. *Weather*, **65**, 180–185,
25 doi:<http://dx.doi.org/10.1002/wea.543>.
- 26 Wilkes, A., A. Reisinger, E. Wollenberg, and S. Van Dijk, 2017: Measurement, reporting and
27 verification of livestock GHG emissions by developing countries in the UNFCCC: current
28 practices and opportunities for improvement. *CCAFS Rep. No. 17*,. www.ccafs.cgiar.org.
- 29 Wilkinson, E., 2018: *Forecasting hazards: averting disasters early action at scale*.
- 30 Wilkinson, E., L. Weingärtner, R. Choularton, M. Bailey, M. Todd, D. Kniveton, and C. C. Venton,
31 2018: *Implementing forecast-based early action at scale*.
- 32 Willemen, L., B. Burkhard, N. Crossman, E. G. Drakou, and I. Palomo, 2015: Editorial: Best practices
33 for mapping ecosystem services. *Ecosystem Services*.
- 34 Willenbockel, D., 2012: *Extreme weather events and crop price spikes in a changing climate.
35 Illustrative global simulation scenarios*.
- 36 Williams, A. P., and J. T. Abatzoglou, 2016: Recent Advances and Remaining Uncertainties in
37 Resolving Past and Future Climate Effects on Global Fire Activity. *Curr. Clim. Chang. Reports*,
38 **2**, 1–14, doi:10.1007/s40641-016-0031-0.
- 39 Williams, B. K., 2011: Adaptive management of natural resources-framework and issues. *J. Environ.
40 Manage.*, **92**, 1346–1353, doi:10.1016/j.jenvman.2010.10.041.
- 41 Williams, D. A., and K. E. Dupuy, 2018: Will REDD+ Safeguards Mitigate Corruption? Qualitative
42 Evidence from Southeast Asia. *J. Dev. Stud.*, 1–16, doi:10.1080/00220388.2018.1510118.
43 <https://www.tandfonline.com/doi/full/10.1080/00220388.2018.1510118> (Accessed November 2,
44 2018).
- 45 Williams, S. E., E. E. Bolitho, and S. Fox, 2003: Climate change in Australian tropical rainforests: an
46 impending environmental catastrophe. *Proc. R. Soc. London. Ser. B Biol. Sci.*, **270**, 1887–1892,
47 doi:10.1098/rspb.2003.2464.

- 1 <http://www.royalsocietypublishing.org/doi/10.1098/rspb.2003.2464> (Accessed April 14, 2019).
- 2 Williamson, T. B., and H. W. Nelson, 2017: Barriers to enhanced and integrated climate change
3 adaptation and mitigation in Canadian forest management. *Can. J. For. Res.*, doi:10.1139/cjfr-
4 2017-0252.
- 5 Wilson, G. L., B. J. Dalzell, D. J. Mulla, T. Dogwiler, and P. M. Porter, 2014: Estimating water
6 quality effects of conservation practices and grazing land use scenarios. *J. Soil Water Conserv.*,
7 doi:10.2489/jswc.69.4.330.
- 8 Wilson, R. S., and Coauthors, 2016: A typology of time-scale mismatches and behavioral
9 interventions to diagnose and solve conservation problems. *Conserv. Biol.*, **30**, 42–49,
10 doi:10.1111/cobi.12632.
- 11 Win, Z. C., N. Mizoue, T. Ota, T. Kajisa, S. Yoshida, T. N. Oo, and H.-O. Ma, 2018: Differences in
12 consumption rates and patterns between firewood and charcoal_ A case study in a rural area of
13 Yedashe Township, Myanmar. *Biomass and Bioenergy*, **109**, 39–46,
14 doi:10.1016/j.biombioe.2017.12.011. <https://doi.org/10.1016/j.biombioe.2017.12.011> (Accessed
15 April 15, 2019).
- 16 Winemiller, K. O., and Coauthors, 2016: DEVELOPMENT AND ENVIRONMENT. Balancing
17 hydropower and biodiversity in the Amazon, Congo, and Mekong. *Science*, **351**, 128–129,
18 doi:10.1126/science.aac7082. <http://www.ncbi.nlm.nih.gov/pubmed/26744397> (Accessed May
19 25, 2018).
- 20 Winickoff, D. E., and M. Mondou, 2017: The problem of epistemic jurisdiction in global governance:
21 The case of sustainability standards for biofuels. *Soc. Stud. Sci.*, **47**, 7–32,
22 doi:10.1177/0306312716667855. <https://doi.org/10.1177/0306312716667855> (Accessed April
23 12, 2019).
- 24 Wise, R. M., I. Fazey, M. Stafford Smith, S. E. Park, H. C. Eakin, E. R. M. Archer Van Garderen, and
25 B. Campbell, 2014: Reconceptualising adaptation to climate change as part of pathways of
26 change and response. *Glob. Environ. Chang.*, **28**, 325–336,
27 doi:10.1016/j.gloenvcha.2013.12.002.
- 28 ———, J. R. A. Butler, W. Suadnya, K. Puspadi, I. Suharto, and T. D. Skewes, 2016: How climate
29 compatible are livelihood adaptation strategies and development programs in rural Indonesia?
30 *Clim. Risk Manag.*, **12**, doi:10.1016/j.crm.2015.11.001.
- 31 Wittmann, M., S. Chandra, K. Boyd, and C. Jerde, 2016: Implementing invasive species control: a
32 case study of multi-jurisdictional coordination at Lake Tahoe, USA. *Manag. Biol. Invasions*,
33 doi:10.3391/mbi.2015.6.4.01.
- 34 Wodon, Q., and H. Zaman, 2010: Higher Food Prices in Sub-Saharan Africa: Poverty Impact and
35 Policy Responses. *World Bank Res. Obs.*, **25**, 157–176, doi:10.1093/wbro/lkp018.
- 36 Woodward, M., Z. Kapelan, and B. Gouldby, 2014: Adaptive flood risk management under climate
37 change uncertainty using real options and optimisation. *Risk Anal.*, **34**.
- 38 Woolf, D., D. Solomon, and J. Lehmann, 2018: Land restoration in food security programmes:
39 synergies with climate change mitigation. *Clim. Policy*, 1–11,
40 doi:10.1080/14693062.2018.1427537.
41 <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1427537> (Accessed May 24,
42 2018).
- 43 Woollen, E., and Coauthors, 2016: Charcoal production in the Mopane woodlands of Mozambique:
44 what are the trade-offs with other ecosystem services? *Philos. Trans. R. Soc. B Biol. Sci.*, **371**,
45 20150315, doi:10.1098/rstb.2015.0315.
46 <http://rstb.royalsocietypublishing.org/lookup/doi/10.1098/rstb.2015.0315> (Accessed May 18,
47 2018).
- 48 World Bank., 2009a: *Environmental crisis or sustainable development opportunity?*

- 1 <http://siteresources.worldbank.org/ezp.sub.su.se/EXTAFRREGTOPENERGY/Resources/71730>
2 [5-1355261747480/World_Bank_Transforming_the_Charcoal_Sector_in_Tanzania.pdf](http://siteresources.worldbank.org/ezp.sub.su.se/EXTAFRREGTOPENERGY/Resources/71730)
3 (Accessed April 5, 2019).
- 4 —, 2009b: *Environmental crisis or sustainable development opportunity?*
- 5 World Bank, 2018: *The State of Social Safety Nets 2018*. Washington, DC.,
6 <https://openknowledge.worldbank.org/handle/10986/29115>.
- 7 World Food Programme, 2018: *Food Security Climate Resilience Facility (FoodSECuRE)*.
- 8 World Health Organization, 2014: *Quantitative risk assessment of the effects of climate change on*
9 *selected causes of death, 2030s and 2050s*. World Health Organization,
10 http://apps.who.int/iris/bitstream/10665/134014/1/9789241507691_eng.pdf (Accessed
11 December 29, 2017).
- 12 Wreford, A., and A. Renwick, 2012: Estimating the costs of climate change adaptation in the
13 agricultural sector. *CAB Rev. Perspect. Agric. Vet. Sci. Nutr. Nat. Resour.*, **7**,
14 doi:10.1079/PAVSNR20127040.
- 15 Wreford, A., A. Ignaciuk, and G. Gruère, 2017: Overcoming barriers to the adoption of climate-
16 friendly practices in agriculture. *OECD Food, Agric. Fish. Pap.*, **101**, doi:10.1787/97767de8-en.
- 17 Wu, X., Y. Lu, S. Zhou, L. Chen, and B. Xu, 2016: Impact of climate change on human infectious
18 diseases: Empirical evidence and human adaptation. *Environ. Int.*, **86**, 14–23,
19 doi:10.1016/J.ENVINT.2015.09.007.
20 <https://www.sciencedirect.com/science/article/pii/S0160412015300489> (Accessed October 1,
21 2018).
- 22 Wunder, S., and R. Bodle, 2019: Achieving land degradation neutrality in Germany: Implementation
23 process and design of a land use change based indicator. *Environ. Sci. Policy*, **92**, 46–55,
24 doi:10.1016/J.ENVSCI.2018.09.022.
25 <https://www.sciencedirect.com/science/article/pii/S146290111830652X> (Accessed April 24,
26 2019).
- 27 Xu, J., E. T. Ma, D. Tashi, Y. Fu, Z. Lu, and D. Melick, 2005: Integrating sacred knowledge for
28 conservation: Cultures and landscapes in Southwest China. *Ecol. Soc.*, doi:10.5751/ES-01413-
29 100207.
- 30 Yamagata, Y., N. Hanasaki, A. Ito, T. Kinoshita, D. Murakami, and Q. Zhou, 2018: Estimating
31 water–food–ecosystem trade-offs for the global negative emission scenario (IPCC-RCP2.6).
32 *Sustain. Sci.*, doi:10.1007/s11625-017-0522-5.
- 33 Yamana, T. K., A. Bomblies, and E. A. B. Eltahir, 2016: Climate change unlikely to increase malaria
34 burden in West Africa. *Nat. Clim. Chang.*, **6**, 1009–1013, doi:10.1038/nclimate3085.
35 <http://www.nature.com/doi/10.1038/nclimate3085> (Accessed December 29, 2017).
- 36 Yami, M., C. Vogl, and M. Hauser, 2009: Comparing the Effectiveness of Informal and Formal
37 Institutions in Sustainable Common Pool Resources Management in Sub-Saharan Africa.
38 *Conserv. Soc.*, **7**, 153, doi:10.4103/0972-4923.64731.
- 39 —, —, and —, 2011: Informal institutions as mechanisms to address challenges in communal
40 grazing land management in Tigray, Ethiopia. *Int. J. Sustain. Dev. World Ecol.*, **18**, 78–87,
41 doi:10.1080/13504509.2010.530124.
- 42 Yang, J., H. Tian, B. Tao, W. Ren, J. Kush, Y. Liu, and Y. Wang, 2014a: Spatial and temporal
43 patterns of global burned area in response to anthropogenic and environmental factors:
44 Reconstructing global fire history for the 20th and early 21st centuries. *J. Geophys. Res.*
45 *Biogeosciences*, **119**, 249–263, doi:10.1002/2013JG002532.
46 <http://doi.wiley.com/10.1002/2013JG002532> (Accessed April 14, 2019).
- 47 —, —, —, —, —, —, and —, 2014b: Spatial and temporal patterns of global burned
48 area in response to anthropogenic and environmental factors: Reconstructing global fire history

- 1 for the 20th and early 21st centuries. *J. Geophys. Res. Biogeosciences*, **119**, 249–263,
2 doi:10.1002/2013JG002532.
- 3 Yang, W., and Q. Lu, 2018: Integrated evaluation of payments for ecosystem services programs in
4 China: a systematic review. *Ecosyst. Heal. Sustain.*, doi:10.1080/20964129.2018.1459867.
- 5 Youn, S.-J., W. W. Taylor, A. J. Lynch, I. G. Cowx, T. D. Beard Jr, D. Bartley, and F. Wu, 2014:
6 Inland capture fishery contributions to global food security and threats to their future. *Glob.*
7 *Food Sec.*, **3**, 142–148.
- 8 Young, H. S., and Coauthors, 2017a: Interacting effects of land use and climate on rodent-borne
9 pathogens in central Kenya. *Philos. Trans. R. Soc. B Biol. Sci.*, **372**, 20160116,
10 doi:10.1098/rstb.2016.0116.
11 <http://rstb.royalsocietypublishing.org/lookup/doi/10.1098/rstb.2016.0116> (Accessed April 23,
12 2019).
- 13 ———, and Coauthors, 2017b: Interacting effects of land use and climate on rodent-borne pathogens in
14 central Kenya. *Philos. Trans. R. Soc. B Biol. Sci.*, doi:10.1098/rstb.2016.0116.
- 15 Young, O. R., 2017a: *Governing Complex Systems. Social Capital for the Anthropocene.*
16 Massachusetts Institute of Technology, Massachusetts,.
- 17 ———, 2017b: Beyond Regulation: Innovative strategies for governing large complex systems.
18 *Sustain.*, doi:10.3390/su9060938.
- 19 Yousefpour, R., and M. Hanewinkel, 2016: Climate Change and Decision-Making Under Uncertainty.
20 *Curr. For. Reports*, **2**, 143–149, doi:10.1007/s40725-016-0035-y.
21 <http://dx.doi.org/10.1007/s40725-016-0035-y>.
- 22 Yumkella, K. K., and P. T. Yillia, 2015: Framing the Water-energy Nexus for the Post-2015
23 Development Agenda. *Aquat. Procedia*, **5**, 8–12,
24 doi:https://doi.org/10.1016/j.aqpro.2015.10.003.
25 <http://www.sciencedirect.com/science/article/pii/S2214241X15002813>.
- 26 Zahawi, R. A., J. L. Reid, and K. D. Holl, 2014: Hidden costs of passive restoration. *Restor. Ecol.*, **22**,
27 284–287.
- 28 Zahran, S., S. D. Brody, W. E. Highfield, and A. Vedlitz, 2010: Non-linear incentives, plan design,
29 and flood mitigation: The case of the Federal Emergency Management Agency’s community
30 rating system. *J. Environ. Plan. Manag.*, doi:10.1080/09640560903529410.
- 31 Zanzanaini, C., B. T. Trần, C. Singh, A. Hart, J. Milder, and F. DeClerck, 2017: Integrated landscape
32 initiatives for agriculture, livelihoods and ecosystem conservation: An assessment of
33 experiences from South and Southeast Asia. *Landsc. Urban Plan.*,
34 doi:10.1016/j.landurbplan.2017.03.010.
- 35 Zarfl, C., A. E. Lumsdon, J. Berlekamp, L. Tydecks, and K. Tockner, 2014: A global boom in
36 hydropower dam construction. *Aquat. Sci.*, **77**, 161–170, doi:10.1007/s00027-014-0377-0.
- 37 ———, ———, ———, ———, and ———, 2015: A global boom in hydropower dam construction. *Aquat.*
38 *Sci.*, **77**, 161–170, doi:10.1007/s00027-014-0377-0. [http://link.springer.com/10.1007/s00027-](http://link.springer.com/10.1007/s00027-014-0377-0)
39 [014-0377-0](http://link.springer.com/10.1007/s00027-014-0377-0) (Accessed May 25, 2018).
- 40 Zeng, Z., J. Liu, P. H. Koeneman, E. Zarate, and A. Y. Hoekstra, 2012: Assessing water footprint at
41 river basin level: a case study for the Heihe River Basin in northwest China. *Hydrol. Earth Syst.*
42 *Sci.*, **16**, 2771–2781, doi:10.5194/hess-16-2771-2012. [https://www.hydrol-earth-syst-](https://www.hydrol-earth-syst-sci.net/16/2771/2012/)
43 [sci.net/16/2771/2012/](https://www.hydrol-earth-syst-sci.net/16/2771/2012/) (Accessed April 14, 2019).
- 44 Zhang, J., C. He, L. Chen, and S. Cao, 2018a: Improving food security in China by taking advantage
45 of marginal and degraded lands. *J. Clean. Prod.*, doi:10.1016/j.jclepro.2017.10.110.
- 46 Zhang, W., T. Zhou, L. Zou, L. Zhang, and X. Chen, 2018b: Reduced exposure to extreme
47 precipitation from 0.5 °C less warming in global land monsoon regions. *Nat. Commun.*, **9**, 3153,

- 1 doi:10.1038/s41467-018-05633-3. <http://www.nature.com/articles/s41467-018-05633-3>
2 (Accessed April 12, 2019).
- 3 Zhao, C., and Coauthors, 2017: Temperature increase reduces global yields of major crops in four
4 independent estimates. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1701762114.
- 5 Zhao, L., M. Oppenheimer, Q. Zhu, J. W. Baldwin, K. L. Ebi, E. Bou-Zeid, K. Guan, and X. Liu,
6 2018: Interactions between urban heat islands and heat waves. *Environ. Res. Lett.*, **13**,
7 doi:10.1088/1748-9326/aa9f73.
- 8 Zheng, H., and Coauthors, 2016: Using ecosystem service trade-offs to inform water conservation
9 policies and management practices. *Front. Ecol. Environ.*, **14**, 527–532.
- 10 Ziv, G., E. Baran, S. Nam, I. Rodríguez-Iturbe, and S. A. Levin, 2012: Trading-off fish biodiversity,
11 food security, and hydropower in the Mekong River Basin. *Proc. Natl. Acad. Sci.*, **109**, 5609–
12 5614.
- 13 Zomer, R. J., A. Trabucco, D. A. Bossio, and L. V. Verchot, 2008: Climate change mitigation: A
14 spatial analysis of global land suitability for clean development mechanism afforestation and
15 reforestation. *Agric. Ecosyst. Environ.*, **126**, 67–80, doi:10.1016/j.agee.2008.01.014.
16 [http://www.cgiar-csi.org/wp-content/uploads/2012/11/Zomer-et-al-2008-A-Spatial-Analysis-of-Global-Land-Suitability-for-Clean-Development-Mechanism-Afforestation-and-](http://www.cgiar-csi.org/wp-content/uploads/2012/11/Zomer-et-al-2008-A-Spatial-Analysis-of-Global-Land-Suitability-for-Clean-Development-Mechanism-Afforestation-and-Reforestation.pdf)
17 [Reforestation.pdf](http://www.cgiar-csi.org/wp-content/uploads/2012/11/Zomer-et-al-2008-A-Spatial-Analysis-of-Global-Land-Suitability-for-Clean-Development-Mechanism-Afforestation-and-Reforestation.pdf) (Accessed May 18, 2018).
- 19 Zoogah, D. B., M. W. Peng, and H. Woldu, 2015: Institutions, Resources, and Organizational
20 Effectiveness in Africa. *Acad. Manag. Perspect.*, **29**, 7–31, doi:10.5465/amp.2012.0033.
- 21 Zulu, L. C., and R. B. Richardson, 2013a: Charcoal, livelihoods, and poverty reduction: Evidence
22 from sub-Saharan Africa. *Energy Sustain. Dev.*, **17**, 127–137, doi:10.1016/j.esd.2012.07.007.
23 <http://dx.doi.org/10.1016/j.esd.2012.07.007> (Accessed April 14, 2019).
- 24 —, and —, 2013b: Charcoal, livelihoods, and poverty reduction: Evidence from sub-Saharan
25 Africa. *Energy Sustain. Dev.*, **17**, 127–137, doi:10.1016/j.esd.2012.07.007.
- 26 ScienceDirect - Articles Related To: How the characteristics of innovations impact their adoption:
27 An exploration of climate-smart agricultural innovations in South Africa Original Research
28 Article Journal of Cleaner Production, Volume 172, 20 January 2018, Pages 3825-3840.
29 [http://www.sciencedirect.com/science?_ob=ArticleListURL&_method=list&_ArticleListID=-](http://www.sciencedirect.com/science?_ob=ArticleListURL&_method=list&_ArticleListID=-1247310770&_sort=v&_st=17&view=c&_origin=related_art&panel=citeRelatedArt&_mlktType=Journal&md5=4504f38f5b3851b78a50c7192af6a30d&searchtype=a)
30 [1247310770&_sort=v&_st=17&view=c&_origin=related_art&panel=citeRelatedArt&_mlktTyp](http://www.sciencedirect.com/science?_ob=ArticleListURL&_method=list&_ArticleListID=-1247310770&_sort=v&_st=17&view=c&_origin=related_art&panel=citeRelatedArt&_mlktType=Journal&md5=4504f38f5b3851b78a50c7192af6a30d&searchtype=a)
31 [e=Journal&md5=4504f38f5b3851b78a50c7192af6a30d&searchtype=a](http://www.sciencedirect.com/science?_ob=ArticleListURL&_method=list&_ArticleListID=-1247310770&_sort=v&_st=17&view=c&_origin=related_art&panel=citeRelatedArt&_mlktType=Journal&md5=4504f38f5b3851b78a50c7192af6a30d&searchtype=a) (Accessed December 29,
32 2017).

33

34 **References to be included in bibliography using Mendeley for the cross chapter box on**
35 **Indigenous and Local Knowledge**

- 36 Altieri, Miguel A., and Clara I. Nicholls. 2017. The adaptation and mitigation potential of traditional
37 agriculture in a changing climate. *Climatic Change* 140.1: 33–45
- 38 Brondizio, Eduardo S., and François-Michel Le Tourneau. 2016. Environmental governance for all.
39 *Science* 352.6291: 1272–1273.
- 40 Cochran, Ferdooz V., Nathaniel A. Bronsell, Aloisio Cabalzar, et al. 2015. Indigenous ecological
41 calendars define scales for climate change and sustainability assessments. *Sustainability Science*
42 11.1: 69–89.
- 43 Ens, Emilie J., Pentina Pert, Philip A. Clarke, et al. 2015. Indigenous biocultural knowledge in
44 ecosystem science and management: Review and insight from Australia. *Biological Conservation*
45 181:133–149.
- 46 Falkowski, Tomasz B., Stewart A. W. Diemont, Adolfo Chankin, and David Douterlungne. 2016.
47 Lacandon Maya traditional ecological knowledge and rainforest restoration: Soil fertility beneath
48 six agroforestry system trees. *Ecological Engineering* 92:210–217.

- 1 Gómez-Baggethun, E., et al. 2010. Traditional Ecological Knowledge Trends in the Transition to a
2 Market Economy: Empirical Study in the Doñana Natural Areas. *Conservation Biology*, 24.
- 3 Janif, Shaiza Z., Patrick D. Nunn, Paul Geraghty, William Aalbersberg, Frank R. Thomas, and
4 Mereoni Camailakeba. 2016. Value of traditional oral narratives in building climate-change
5 resilience: Insights from rural communities in Fiji. *Ecology and Society* 21.2: art7.
- 6 Kimmerer, Robin Wall. 2000. Native knowledges for native ecosystems. *Journal of Forestry* 98.8: 4–
7 9.
- 8 Krätli, S. (2008). Cattle breeding, complexity and mobility in a structurally unpredictable
9 environment: the WoDaaBe herders of Niger. *Nomadic Peoples*, 12(1), 11-41.
- 10 McCarter, Joe, et al. “The Challenges of Maintaining Indigenous Ecological Knowledge.” *Ecology*
11 *and Society*, vol. 19, no. 3, 2014.
- 12 Myers, Fred. 1991. *Pintupi country, Pintupi self: Sentiment, place, and politics among western desert*
13 *aborigines*. Berkeley: Univ. of California Press
- 14 Orlove, Ben, Carla Roncoli, Merit R. Kabugo, and Abushen Majugu. 2010. Indigenous climate
15 knowledge in southern Uganda: The multiple components of a dynamic regional system. *Climatic*
16 *Change* 100.2: 243–265.
- 17 Oteros-Rozas, E., et al. 2013. Traditional Ecological Knowledge Among Transhumant Pastoralists in
18 Mediterranean Spain: Learning for Adaptation to Global Change. *Ecology and Society* 18 (3).
- 19 Pungetti, Gloria, Gonzalo Oviedo, and Della Hooke. 2012. *Sacred species and sites: Advances in*
20 *biocultural conservation*. Cambridge, UK: Cambridge Univ. Press
- 21 Richards, Paul. 1985. *Indigenous agricultural revolution: Ecology and food production in West*
22 *Africa*. Boulder, CO: Westview
- 23 Spoon, Jeremy. “Quantitative, Qualitative, and Collaborative Methods: Approaching Indigenous
24 Ecological Knowledge Heterogeneity.” *Ecology and Society* 19, no. 3 (2014): art33.
25 <http://www.ecologyandsociety.org/vol19/iss3/art33/>.
- 26 Stringer L, Dyer J, Reed MS, Dougill AJ, Twyman C, Mkwambisi D. 2009. Adaptations to climate
27 change, drought and desertification: insights to enhance policy in southern Africa. *Environmental*
28 *Science and Policy*. 12(7), pp. 748-765.
- 29 Stringer LC, Quinn CH, Le HTV, Msuya F, Pezzuti J, Dallimer M, Afionis S, Berman R, Orchard SE,
30 Rijal M. 2018. A new framework to enable equitable outcomes: resilience and nexus approaches
31 combined. *Earth's Future*. 6(6), pp. 902-918
- 32 Tengö, Maria, Eduardo S. Brondizio, Thomas Elmqvist, Pernilla Maler, and Marja Spierenburg. 2014.
33 Connecting diverse knowledge systems for enhanced ecosystem governance: The multiple evidence
34 base approach. *Ambio* 43.5: 579–591.
- 35 Usher, Peter J. 2000. Traditional ecological knowledge in environmental assessment and
36 management. *Arctic* 53.2: 183–193
- 37 White, J., & Morton, J. (2005). Mitigating impacts of HIV/AIDS on rural livelihoods: NGO
38 experiences in sub-Saharan Africa. *Development in Practice*, 15(2), 186-199.
- 39

1 **Supplementary Material**

2 Additional material on Section 7.2.2 in separate file.

3 **Additional material from Section 7.2.4:**

4

5 Table 7.1 Appendix

| Land-Climate-Society interaction Hazard | Exposure | Vulnerability | Risk | Policy Response (Indicative) | References |
|---|------------------------|--|---|--|--|
| Inefficient carbon capture and storage | Global | Developing countries | <ul style="list-style-type: none"> • Disincentivising low carbon pathways/renewables • Loss of water resources, biodiversity and ecosystem services • Dangerous climate change ie SSP2 and SSP3 pathways | <ul style="list-style-type: none"> • Certification • Transdisciplinary research on feasibility and pilot projects | (Smith et al. 2016; Fuss et al. 2014; Torvanger 2019b) |
| Increasing incidences of wildfires at the wildland-urban interface | USA, Canada, Australia | Peri-urban communities next to forests | <ul style="list-style-type: none"> • Loss of life and property | <ul style="list-style-type: none"> • Willingness to pay for prescribed fire • Local early warning and communication • Wildlife frequency and risk mapping | (Abatzoglou and Williams 2016; Gan et al. 2014; Kaval et al. 2007; Mozumder et al. 2009; Brenkert–Smith et al. 2006)(Radeloff et al. 2018) |
| | | | <ul style="list-style-type: none"> • • | <ul style="list-style-type: none"> • • | |

| Land-Climate-Society interaction Hazard | Exposure | Vulnerability | Risk | Policy Response (Indicative) | References |
|---|---|---|---|--|---|
| Use of land for renewable energy | India, China, semi-arid regions | Pastoralists Farmers Biodiversity | <ul style="list-style-type: none"> • Loss of biodiversity and ecosystem services • • | <ul style="list-style-type: none"> • See 7.5.6 • • | See 7.5.6 |
| Urban air pollution from surrounding land-use | Urban centres existing and emerging in developing countries | Marginalized communities, pedestrians, commuters, street vendors, children | <ul style="list-style-type: none"> • Health risk • allergic respiratory diseases | <ul style="list-style-type: none"> • Air pollution regulation • Fuel conversion to clean energy • Incentives to reduce crop stubble burning | (Sharma et al., 2013, D'Amato et al., 2010) |
| Severe weather hazards for cultural heritage (sensitive historic material) | Regions with increase precipitation Increase in the freeze-thaw cycle in northern regions Extreme heat and drought in dry area Landslide and | Buildings and sites in areas with increasing intensities of rain and humidity | <ul style="list-style-type: none"> • Loss of culture and identity | <ul style="list-style-type: none"> • Restoration and protection measures incorporated in regulations and management plan | (Sesana et al, 2018, Sabbioni et al., 2008) |

| Land-Climate-Society interaction Hazard | Exposure | Vulnerability | Risk | Policy Response (Indicative) | References |
|---|-------------------------|---------------|------|---------------------------------|------------|
| | groundwater flooding | | | | |

1

1 Supplementary Material

3 Supplementary information to Section 7.2

5 The burning embers diagrams (Figure 7.1, 7.2 and 7.3) outline risks associated with climate
6 change as a function of global warming, socio-economic development and mitigation choices.
7 Diagrams indicate transitions between undetectable, moderate, high, and very high risks to
8 humans and ecosystems. The method is based on a literature review of estimated impacts at
9 different global mean surface temperature levels (O’Neill et al. 2017) on different components of
10 desertification, land degradation and food security, including emerging literature on Shared
11 Socio-economic Pathways (SSPs) as well as literature from IPCC AR5 and SR15.

13 Most studies focus on changes in hazards as a function of climate change (e.g. as represented by
14 RCP scenarios or other climate change scenarios) or climate change superimposed on present-day
15 exposure. Only a limited number of studies focus on changes in risk as a function of both RCPs
16 and SSPs (climate and socio-economic change and adaptation decisions). This was addressed by
17 splitting the embers into different figures. Figure 7.1 focuses on the impact of climate change on
18 risk, under present-day exposure and vulnerability. Figure 7.2 examines the relationship between
19 climate change and risks under two SSPs (SSP1 and SSP3). Figure 7.3 depicts risks to humans
20 and ecosystems as a function of the land area employed for mitigation through bioenergy
21 plantations.

23 Further, a formal expert elicitation protocol, based on the modified-Delphi technique (Mukherjee
24 et al. 2015) and the Sheffield Elicitation Framework (Oakley and O’Hagan 2016; Gosling 2018),
25 was followed to develop threshold judgments on risk transitions. Specifically, experts participated
26 in a multi-round elicitation process, with feedback of group opinion provided after each round:
27 the first two rounds involved independent anonymous threshold judgment, and the final round
28 involved a group consensus discussion (von der Gracht 2012). To strengthen the rigor of
29 developing expert consensus on risk transitions (Hasson and Keeney 2011), the protocol pre-
30 specified the following prior to beginning the elicitation exercise (Grant et al. 2018): the research
31 question, eligibility criteria and strategy to recruit experts, research materials, data collection
32 procedure, and analysis plan. This systematic process of developing expert consensus on
33 threshold judgments for risk transitions can better inform subsequent analytical approaches—an
34 approach that may be further developed for use in future IPCC cycles (Bojke et al. 2010; Sperber
35 et al. 2013). References for the current and past assessments are listed at the end of this document
36 and by the relevant tables.

38 **Table SM7.1: literature considered in the expert judgement of risk transitions for figure 7.1**

| Reference | Risk | variable (unit) | Direction of impact | climate scenario | Time frame | D/A of current impact | Impact at 1 degree | Impact at 2 degree | Impact at 3 degree | Impact at 4 degree | Impact at 4.5 degree | Adaptation potential | Region (Including Regional Differences) |
|--|--------------------|-----------------|-----------------------------------|------------------|------------|-----------------------|----------------------------|--------------------------|-------------------------|----------------------|----------------------|-------------------------|---|
| AVAILABILITY | | | | | | | | | | | | | |
| Rosenzweig, Cynthia, Joshua Elliott, Delphine Deryng, Alex C. Ruane, Christoph Müller, Almut | Availability Yield | yield | Strong negative effect on yields, | NA | | - | See Figure 1. Maize mid to | Maize - 20 to +5 % yeild | Maize about - 20 to +5% | Maize - +15 to minus | Maize is now all | Between 3 and 4 degrees | Use RCPs so could examine yield |

| | | | | | | | | | | | | | |
|--|---|--|--|--------|------|---|--|--------|---|--|--------------------------|--|----------------------------------|
| Arnoeth, Kenneth J. Boote, et al. 2014. "Assessing Agricultural Risks of Climate Change in the 21st Century in a Global Gridded Crop Model Intercomparison." Proceedings of the National Academy of Sciences. https://doi.org/10.1073/pnas.1222463110 . | | | especiall y at higher levels of warming and at lower latitudes, | | | | high latitude is -10 to +15 % yield change | change | yield change in mid latitude and ALL negative in low latitude | 20% yield change in mid latitude. Catastrophic in low latitude with - 10 to - 60 Percent change! | negative in mid latitude | seems to me catastrophic in low latitude s for maize, wheat also significant declines around 4 degrees and same for rice | according to different pathways. |
| Zscheischler, Jakob, Seth Westra, Bart J.J.M. Van Den Hurk, Sonia I. Seneviratne, Philip J. Ward, Andy Pitman, Amir Aghakouchak, et al. 2018. "Future Climate Risk from Compound Events." Nature Climate Change. https://doi.org/10.1038/s41558-018-0156-3 . | Availability (crop failure) | crop yield | " increases the likelihood of such events considerably, and may make events of the rarity of the Russian event foreseeable and to some extent predictable" | Review | 2010 | - | - | - | - | - | - | - | |
| IPCC Special Report on Global Warming of 1.5°C, 2018 | Availability (crop yields) | yield | Decrease to yields | NA | | - | - | - | - | - | - | Limitin g global warmin g to 1.5°C compar ed to 2°C would result in a lower global reducti on in crop yields | |
| Medina, Angel, Asya Akbar, Alaa Baazeem, Alicia Rodriguez, and Naresh Magan. 2017. "Climate Change, Food Security and Mycotoxins: Do We Know Enough?" Fungal Biology Reviews. https://doi.org/10.1016/j.fbr.2017.04.002 . | Availability (increased loss of crops and livestock ; increased pest burden, increased disease burden; higher post-harvest losses due to mycotox ins) | infection of staple food commodities by fungal diseases pre-harvest and by spoilage fungi post-harvest | reduced availability of food | NA | | - | - | - | - | - | - | | low to moderate |

| | | | | | | | | | | | | |
|---|---|----------------------------|-------------------------------------|-----------|-----------|----------|----------|----------|----------|----------|--|--|
| <p>Paterson, R. R.M., and N. Lima. 2011. "Further Mycotoxin Effects from Climate Change." Food Research International. https://doi.org/10.1016/j.foodres.2011.05.038.</p> | <p>Availability (increased loss of crops and livestock ; increased pest burden, increased disease burden; higher post-harvest losses due to mycotoxins)</p> | <p>crops after harvest</p> | <p>reduced availability of food</p> | <p>NA</p> | <p>NA</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>unclear. "Crops introduced to exploit altered climate may be subject to fewer mycotoxin producing fungi (the "Parasites Lost" phenomenon). Increased mycotoxins and UV radiation may cause fungi to mutate on crops and produce different mycotoxins"</p> |
| <p>Magan, N., A. Medina, and D. Aldred. 2011. "Possible Climate-Change Effects on Mycotoxin Contamination of Food Crops Pre- and Postharvest." Plant Pathology. https://doi.org/10.1111/j.1365-3059.2010.02412.x.</p> | <p>Availability (increased loss of crops and livestock ; increased pest burden, increased disease burden; higher post-harvest losses due to mycotoxins)</p> | <p>crops after harvest</p> | <p>reduced availability of food</p> | <p>NA</p> | <p>NA</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>from high risk to permanent between 3 and 5 degrees</p> | <p>low to moderate</p> |
| <p>Rivera-Ferre, M. G., M. Di Masso, I. Vara, M. Cuellar, A. Calle, M. Mailhos, F. López-i-Gelats, G. Bhatta, and D. Gallar. 2016. "Local Agriculture Traditional Knowledge to Ensure Food Availability in a Changing Climate: Revisiting Water Management Practices in the Indo-Gangetic Plains." Agroecology and Sustainable Food Systems. https://doi.org/10.1080/21683565.2016.1215368.</p> | <p>Availability (increased loss of crops and livestock ; increased pest burden, increased disease burden; higher post-harvest losses due to mycotoxins)</p> | <p>crop yield</p> | <p>reduced availability of food</p> | <p>NA</p> | <p>NA</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>Local/traditional knowledge in agriculture (LTKA) is proposed in this article as valid knowledge to ensure food availability under climate change, given its long experience in dealing with climate variability</p> |

| | | | | | | | | | | | | | | |
|---|--|-----------------------|------------------|---|---|---|---|--|---|---------------------------------------|---|---|---|--|
| Zimmermann, Andrea, Heidi Webber, Gang Zhao, Frank Ewert, Johannes Kros, Joost Wolf, Wolfgang Britz, and Wim de Vries. 2017. "Climate Change Impacts on Crop Yields, Land Use and Environment in Response to Crop Sowing Dates and Thermal Time Requirements." Agricultural Systems. https://doi.org/10.1016/j.agsy.2017.07.007 . | Availability (increased yields if management assumptions hold, thermal management) | crop yields in Europe | increased yields | three SRES climate change scenarios to 2050 | three SRES climate change scenarios to 2050 | - | - | - | - | - | - | - | high | |
| Faye, Babacar, Heidi Webber, Jesse B. Naab, Dilys S. MacCarthy, Myriam Adam, Frank Ewert, John P.A. Lamers, et al. 2018. "Impacts of 1.5 versus 2.0 °c on Cereal Yields in the West African Sudan Savanna." Environmental Research Letters. https://doi.org/10.1088/1748-9326/aaab40 . | Availability (modeled crop yield) | crop yield | negative | NA | | - | - | - | - | - | - | - | between 1 and 2 with success of intensification the key factor making the difference between whether risk remains moderate or red to purple | low to moderate ("despite the larger losses, yields were always two to three times higher with intensification, irrespective of warming scenario") |
| Tesfaye, Kindie, P. H. Zaidi, Sika Gbegbelegbe, Christian Boeber, Dil Bahadur Rahut, Fite Getaneh, K. Seetharam, Olaf Erenstein, and Clare Stirling. 2017. "Climate Change Impacts and Potential Benefits of Heat-Tolerant Maize in South Asia." Theoretical and Applied Climatology. https://doi.org/10.1007/s00704-016-1931-6 . | Availability (modeled crop yield) | crop yield | negative | NA | | - | - | "at regional scale, they found maize yields declines in 2050 of up to 12% to 14% in rainfed and irrigated maize" | - | - | - | - | between 1.0 and 1.5 | low |
| Scheelbeek, Pauline F. D., Frances A. Bird, Hanna L. Tuomisto, Rosemary Green, Francesca B. Harris, Edward J. M. Joy, Zaid Chalabi, Elizabeth Allen, Andy Haines, and Alan D. Dangour. 2018. "Effect of Environmental Changes on Vegetable and Legume Yields and Nutritional Quality." Proceedings of the National Academy of Sciences. https://doi.org/10.1073/pnas.1800442115 . | Availability (modeled crop yield) | crop yield | negative | NA | | - | - | - | - | mean yield declines of fruits - 31.5% | - | - | | |
| Rippke, Ulrike, Julian Ramirez-Villegas, Andy | Availability | crop yield | negative | NA | to end | - | - | "30-60% of | - | - | - | - | between 2.6 | low |

| | | | | | | | | | | | | | |
|--|--|------------|--|-------------------|--|---|---|---|---|---|---|--|--------|
| <p>Jarvis, Sonja J. Vermeulen, Louis Parker, Flora Mer, Bernd Diekkruiger, Andrew J. Challinor, and Mark Howden. 2016. "Timescales of Transformational Climate Change Adaptation in Sub-Saharan African Agriculture." <i>Nature Climate Change</i>. https://doi.org/10.1038/nclimate2947.</p> | (modeled crop yield) | | | | of 21st century | | | common bean growing area and 20-40% of banana growing areas in Africa will lose viability in 2078-2098 with a global temperature increase of 2.6 and 4.0" | | | | and 4.0 ("30-60% of common bean growing area and 20-40% of banana growing areas in Africa will lose viability in 2078-2098 with a global temperature increase of 2.6 and 4.0") | |
| <p>Bisbis, M. B., N. Gruda, and M. Blanke. 2018: Potential impacts of climate change on vegetable production and product quality - A review. <i>J. Clean. Prod.</i>, 170, 1602–1620, doi:10.1016/j.jclepro.2017.09.224.</p> | Availability (modeled fruit crop yield), and utilization (reduced quality, more spoilage, reduced nutrition) | crop yield | negative | NA | | - | - | - | - | - | - | between 1.0 and 1.5 | medium |
| <p>Tibaldi, Claudia, and David Lobell. 2018. "Estimated Impacts of Emission Reductions on Wheat and Maize Crops." <i>Climatic Change</i>. https://doi.org/10.1007/s10584-015-1537-5.</p> | Availability (models relation between climate variables, CO2 concentrations, and yields) | crop yield | negative | RCP4.5 and RCP8.5 | short (2021 – 2040), medium (2041 – 2060) and long (2061 – 2080) time horizons | - | - | "critical or "lethal" heat extreme | - | - | - | modeling results in RCP8.5 (tripling of lethal heat extremes), modeling results in RCP4.5 (doubling of lethal heat extremes) towards end of 21st century | low |
| <p>Schleussner, Carl Friedrich, Delphine Deryng, Christoph Müller, Joshua Elliott, Fahad Saeed, Christian Folberth, Wenfeng Liu, et al. 2018. "Crop Productivity Changes in 1.5 °c and 2 °c Worlds under Climate Sensitivity Uncertainty."</p> | Availability (reduced yields and soil fertility and increased land degradation for | yield | negative for half a degree additional warming (1.5 to 2) | HAPPI | | - | - | "half a degree warming will also lead to more extreme low yields, in | - | - | - | | |

| | | | | | | | | | | | | | | |
|--|--|-------------------------|--|----|--|---|--------------------|-----------------------------------|---------------------|---------------------|---------------------|--|--|--|
| Environmental Research Letters. https://doi.org/10.1088/1748-9326/aab63b . | some regions and crops) | | | | | | | particular over tropical regions" | | | | | | |
| Ovalle-Rivera, Oriana, Peter Läderach, Christian Bunn, Michael Obersteiner, and Götz Schroth. 2015. "Projected Shifts in Coffea Arabica Suitability among Major Global Producing Regions Due to Climate Change." PLoS ONE. https://doi.org/10.1371/journal.pone.0124155 . | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | yield | Decrease in coffee yields | NA | | - | - | - | - | - | - | | | |
| Bunn, Christian, Peter Läderach, Oriana Ovalle Rivera, and Dieter Kirschke. 2015. "A Bitter Cup: Climate Change Profile of Global Production of Arabica and Robusta Coffee." Climatic Change. https://doi.org/10.1007/s10584-014-1306-x . | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | yield | Decrease in coffee yields by 50% | NA | | - | - | - | - | - | - | | | |
| Roberts, Michael J., and Wolfram Schlenker. 2013. "Identifying Supply and Demand Elasticities of Agricultural Commodities: Implications for the US Ethanol Mandate." American Economic Review. https://doi.org/10.1257/aer.103.6.2265 . 2009 | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | yield | productivity of major crops will decline as a result of climate change, particularly from increasing warming | NA | | - | - | - | - | - | - | | | |
| Peng, S., J. Huang, J. E. Sheehy, R. C. Laza, R. M. Visperas, X. Zhong, G. S. Centeno, G. S. Khush, and K. G. Cassman. 2004. "Rice Yields Decline with Higher Night Temperature from Global Warming." Proceedings of the National Academy of Sciences. https://doi.org/10.1073/pnas.0403720101 . | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | grain yields | Grain yield of rice declined 10% for each 1°C increase in nighttime temperature during the dry season | NA | | - | -10% | -20% | -30% | -40% | -50% | | | |
| Asseng, S., F. Ewert, P. Martre, R. P. Rötter, D. B. Lobell, D. Cammarano, B. A. Kimball, et al. 2015. "Rising Temperatures Reduce Global Wheat Production." Nature Climate Change. https://doi.org/10.1038/nclimate2470 . et al., 2015 | Availability (reduced yields and soil fertility and increased land degradation for some | soy bean & maize yields | while maize and soy bean yields are expected to decline by 6% for each day | NA | | - | -6%/day above 30°C | -12%/day above 30°C | -18%/day above 30°C | -24%/day above 30°C | -30%/day above 30°C | | | |

| | | | | | | | | | | | | | |
|---|--|-----------------------|---|----|----|--|-------|--------|------------------------|--------|------|---|-----------------------------------|
| | regions and crops) | | above 30°C. | | | | | | | | | | |
| Asseng, Senthod, Davide Cammarano, Bruno Basso, Uran Chung, Phillip D. Alderman, Kai Sonder, Matthew Reynolds, and David B. Lobell. 2017. "Hot Spots of Wheat Yield Decline with Rising Temperatures." <i>Global Change Biology</i> . https://doi.org/10.1111/gcb.13530 . | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | wheat yields | wheat yields are expected to decline by 6% for each 1°C increase; | NA | | warmin g is already slowing yield gains at a majorit y of wheat-growin g location s. | -0.06 | -0.12 | -0.18 | -0.24 | -0.3 | tiping point above 28 degrees C, no yield | medium |
| Porter, John R., Liyong Xie, Andrew J Challinor, Kevern Cochrane, S. Mark Howden, Muhammad Mohsin Iqbal, David B. Lobell, and Maria Isabel Travasso. 2014. "Food Security and Food Production Systems." In <i>Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change</i> , 485–533. https://doi.org/10.1111/j.1728-4457.2009.00312.x . | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | crop yields all crops | If global temperature increases beyond 3°C it will have negative yield impacts on all crops | NA | | - | - | - | negati ve yield impact | - | - | | |
| Schleussner, Carl Friedrich, Tabea K. Lissner, Erich M. Fischer, Jan Wohland, Mahé Perrette, Antonius Golly, Joeri Rogelj, et al. 2016. "Differential Climate Impacts for Policy-Relevant Limits to Global Warming: The Case of 1.5 °c and 2 °c." <i>Earth System Dynamics</i> . https://doi.org/10.5194/esd-7-327-2016 . | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | competiti on for land | increasin g competit ion for land from the expansio n of bioenerg y | NA | | - | - | - | - | - | - | | |
| Fischer, Günther, Mahendra Shah, Francesco N. Tubiello, and Harrij Van Velhuizen. 2005. "Socio-Economic and Climate Change Impacts on Agriculture: An Integrated Assessment, 1990-2080." In <i>Philosophical Transactions of the Royal Society B: Biological Sciences</i> . https://doi.org/10.1098/rstb.2005.1744 . | Availability (reduced yields and soil fertility and increased land degradation for some regions and crops) | | Decrease in yields | NA | NA | - | 10% | 10-20% | 10-20% | 10-20% | - | | on-farm and via market mechanisms |
| Smith, Pete, R. Stuart Haszeldine, and Stephen M. Smith. 2016. "Preliminary Assessment of the Potential for, and Limitations to, Terrestrial Negative | Availability (reduced yields and soil fertility and | soil | reduced yields | NA | NA | - | - | - | - | - | - | | moderate |

| | | | | | | | | | | | | | | |
|---|---|------------|--------------------|----|------------------------|---|--------|--------|--------|--------|---|---|----------------------------|-----------------|
| Emission Technologies in the UK.” Environmental Science: Processes and Impacts. https://doi.org/10.1039/c6em00386a . | increase d land degradat ion for some regions and crops) | | | | | | | | | | | | | |
| Challinor, A. J., J. Watson, D. B. Lobell, S. M. Howden, D. R. Smith, and N. Chhetri. 2014. “A Meta-Analysis of Crop Yield under Climate Change and Adaptation.” Nature Climate Change. https://doi.org/10.1038/nclimate2153 . | Availability (reduced yields and soil fertility and increase d land degradat ion for some regions and crops) | crop yield | reduced yields | NA | 2050 to end of century | - | - | - | - | - | - | - | likely between 1.5 and 2.0 | low to moderate |
| FAO 2018a | Availability (reduced yields and soil fertility and increase d land degradat ion for some regions and crops) | crop yield | reduced yields | NA | | - | - | - | - | - | - | - | likely between 1.0 and 1.5 | low to moderate |
| Roberts, Michael J., and Wolfram Schlenker. 2013. “Identifying Supply and Demand Elasticities of Agricultural Commodities: Implications for the US Ethanol Mandate.” American Economic Review. https://doi.org/10.1257/aer.103.6.2265 . 2009 | Availability (reduced yields and soil fertility and increase d land degradat ion for some regions and crops)(3 crops) | | Decrease in yields | NA | | - | 30-46% | 30-46% | 63-80% | 63-80% | - | | | |
| Richard A Betts, Lorenzo Alfieri, John Caesar, Luc Feyen, Laila Gohar, Aristeidis Koutroulis, et al. 2018. “Subject Areas : Author for Correspondence : Changes in Climate Extremes , Fresh Water Availability and Vulnerability to Food Insecurity Projected at 1 . 5 ° C and 2 ° C Global Warming with a Higher-Resolution Global Climate Model.” et al, 2018 | Availability (reduced yields and soil fertility and increase d land degradat ion for some regions and crops)(fo od crops) | yield | decreae | NA | | - | - | - | - | - | - | | | |
| Tigchelaar, M, D Battisti, R.L Naylor, and D.K Ray. 2018. “Probability of Globally Synchronized Maize Production Shocks.” Proceedings of the National Academy of | Availability (reduced yields and soil fertility and increase | | Decrease in yields | NA | | - | - | 7-10% | - | 87% | - | | | |

| | | | | | | | | | | | | | |
|--|--|------------|---|--|-----------|---|---|---|---|---|---|--|---|
| Sciences 115 (26): 6644–49. | d land degradat ion for some regions and crops)(Maize) | | | | | | | | | | | | |
| Leng, Guoyong, and Jim Hall. 2019. “Crop Yield Sensitivity of Global Major Agricultural Countries to Droughts and the Projected Changes in the Future.” Science of the Total Environment. https://doi.org/10.1016/j.scitotenv.2018.10.434 . | Availabi lity (reduced yields and soil fertility and increase d land degradat ion for some regions and crops)(si x crops) | | Declinin g yield (but varies between crops and regions) | NA | | - | - | - | - | - | - | | Study doesn't consider adaptatio ns |
| Bocchiola, D., L. Brunetti, A. Soncini, F. Polinelli, and M. Gianinetto. 2019. “Impact of Climate Change on Agricultural Productivity and Food Security in the Himalayas: A Case Study in Nepal.” Agricultural Systems. https://doi.org/10.1016/j.agsy.2019.01.008 . | Availabi lity (reduced yields and soil fertility and increase d land degradat ion for some regions and crops)(w heat, rice, maize) | | Declinin g | NA | | - | - | - | - | - | - | | Increasing altitude - increases yield for maize and rice slightly |
| Rozenzweig et al. 2017 | Availabi lity (simulat ed wheat and maize yield changes) | crop yield | negative | AgMIP coordinate d global and regional assessment (CGRA) | | - | - | - | - | - | - | betwe n 1.5 and 2.0 | low |
| Parkes et al. 2017 | Availabi lity (simulat ed wheat and maize yield changes) | crop yield | negative | NA | | - | - | - | - | - | - | betwe n 1.0 and 1.5 | low |
| Lombardozi, Danica L., Nicholas G. Smith, Susan J. Cheng, Jeffrey S. Dukes, Thomas D. Sharkey, Alistair Rogers, Rosie Fisher, and Gordon B. Bonan. 2018. “Triose Phosphate Limitation in Photosynthesis Models Reduces Leaf Photosynthesis and Global Terrestrial Carbon Storage.” Environmental Research Letters. https://doi.org/10.1088/1748-9326/aacf68 . | Availabi lity (Yield) | yield | positive effect of CO2 on future crop yields muted by negative impacts of climate | CESM/CL M4.5 under RCP8.5 | 2006-2100 | - | - | - | - | - | - | Corn: -10 to +20% Wheat +40 to +100 %; Soy -10 to +5 %; Rice +10 to +50% | |
| Chen, Y. et al. (2018) ESD Impacts of climate change and climate | Availabi lity (Yield) | yield | decrease in organic | NA | | - | - | - | - | - | - | | |

| | | | | | | | | | | | | | | |
|---|-----------------------------------|--------------------------------|--|----|--|---|--|---|------|------|---|---|--|--|
| extremes on major crops productivity in China at a global warming of 1.5 and 2.0C | | | matter in soil, soil erosion | | | | | | | | | | | |
| Leng, G. (2018) SOTE Keeping global warming within 1.5C reduces future risk of yield loss in the United States: A probabilistic modeling approach | Availability (Yield) | yield | | NA | | - | - | - | - | - | - | | | |
| Byers, Edward, Matthew Gidden, David Leclère, Juraj Balkovic, Peter Burek, Kristie Ebi, Peter Greve, et al. 2018. "Global Exposure and Vulnerability to Multi-Sector Development and Climate Change Hotspots." Environmental Research Letters 13 (5): 055012. https://doi.org/10.1088/1748-9326/aabf45 . | Availability (Yield) | yield | | NA | | - | - | - | - | - | - | | | |
| Xie, Wei, Wei Xiong, Jie Pan, Tariq Ali, Qi Cui, Dabo Guan, Jing Meng, Nathaniel D. Mueller, Erda Lin, and Steven J. Davis. 2018. "Decreases in Global Beer Supply Due to Extreme Drought and Heat." Nature Plants. https://doi.org/10.1038/s41477-018-0263-1 . | Availability barley yields (beer) | yield | Decrease in barley yield, consumption (and hence global beer supply) | NA | | - | - | -3% | -10% | -17% | - | | | |
| Leng, Guoyong, and Jim Hall. 2019. "Crop Yield Sensitivity of Global Major Agricultural Countries to Droughts and the Projected Changes in the Future." Science of the Total Environment. https://doi.org/10.1016/j.scitotenv.2018.10.434 . | Availability Corn Yields | yield | Decrease to yields. | NA | | 2.5% decrease of corn yield for the historical period, which is reduced to 1.8% if accounting for the effects of corn growing pattern changes | Negative corn yield response to warmer growing season, largest yield reduction up to 20% by 1° increase of temperature | majority of impacts will be driven by trends in temperature rather than precipitation | - | - | - | Negative corn yield response to warmer growing season | Corn yield is predicted to decrease by 20~40% by 2050s | |
| Leng, Guoyong. 2018. "Keeping Global Warming within 1.5 °C Reduces Future Risk of Yield Loss in the United States: A Probabilistic Modeling Approach." Science of the Total Environment. https://doi.org/10.1016/j.scitotenv.2018.06.344 . | Availability crop yields | yield | Decrease in yields | NA | | - | - | - | - | - | - | | | |
| Su, B. et al. (2018) Drought losses in China might double between the 1.5C and 2.0C warming. PNAS | Availability crop yields | yield | Decrease in yields | NA | | - | - | - | - | - | - | | | |
| Zhao, Chuang, Bing Liu, Shilong Piao, Xuhui Wang, David B. Lobell, Yao Huang, Mengtian | Availability maize yields | yield, production/ per hectare | Decrease in yield | NA | | - | - | - | - | - | - | | | |

| | | | | | | | | | | | | | |
|---|--|----------------|-------------------------|----|--|---|---|---|-----|------|---|----------------|---|
| Huang, et al. 2017. "Temperature Increase Reduces Global Yields of Major Crops in Four Independent Estimates." Proceedings of the National Academy of Sciences. https://doi.org/10.1073/pnas.1701762114 . | | | | | | | | | | | | | |
| Brisson, Nadine, Philippe Gate, David Gouache, Gilles Charmet, François Xavier Oury, and Frédéric Huard. 2010. "Why Are Wheat Yields Stagnating in Europe? A Comprehensive Data Analysis for France." Field Crops Research. https://doi.org/10.1016/j.fcr.2010.07.012 . | Availability Yield | yield | yield losses/plateauing | NA | | - | - | - | - | - | - | | |
| Lin, M., and P. Huybers. 2012. "Reckoning Wheat Yield Trends." Environmental Research Letters. https://doi.org/10.1088/1748-9326/7/2/024016 . | Availability Yield | yield | yield losses/plateauing | NA | | - | - | - | - | - | - | | |
| Grassini, Patricio, Kent M. Eskridge, and Kenneth G. Cassman. 2013. "Distinguishing between Yield Advances and Yield Plateaus in Historical Crop Production Trends." Nature Communications. https://doi.org/10.1038/ncomms3918 . | Availability Yield | yield | yield losses/plateauing | NA | | - | - | - | - | - | - | | |
| Myers, S.S.; M. R. Smith, S. Guth, C. D. Golden, B. Vaitla, N. D. Mueller, A. D. Dangour, and P. Huybers, 2017: Climate Change and Global Food Systems: Potential Impacts on Food Security and Undernutrition. Annu. Rev. Public Health, 38, 259–277, doi:10.1146/annurev-publhealth-031816-044356. http://www.annualreviews.org/doi/10.1146/annurev-publhealth-031816-044356 . | Availability yield declines | yield | | NA | | - | - | - | - | - | - | | adaptation could lead to crop yields that are 7-15% higher. Gains will be highest in temperate areas but will be unlikely to help tropical maize and wheat production |
| Hasegawa, Tomoko, Shinichiro Fujimori, Petr Havlík, Hugo Valin, Benjamin Leon Bodirsky, Jonathan C. Doelman, Thomas Fellmann, et al. 2018. "Risk of Increased Food Insecurity under Stringent Global Climate Change Mitigation Policy." Nature Climate Change 8 (8): 699–703. https://doi.org/10.1038/s41558-018-0230-x . | Mitigation policy combined with climate effect on yields | available land | | NA | | - | - | - | - | - | - | | |
| ACCESS | | | | | | | | | | | | | |
| Schmidhuber, J., and F. N. Tubiello. 2007. | Access Price | Price | increase in price | NA | | - | - | - | 80% | 170% | - | current period | |

| | | | | | | | | | | | | | |
|---|--|---|------------------------|----|--|-------|--------|--------|--------|--------|--------|------------|---|
| “Global Food Security under Climate Change.” Proceedings of the National Academy of Sciences. https://doi.org/10.1073/pnas.0701976104 , 2007 | (cereal) | | | | | | | | | | | (timewise) | |
| IPCC AR4 (Easterling et al, 2007) | Access Price (cereal) | price | increase in price | NA | | - | 10-30% | 10-30% | 10-40% | 10-40% | 10-40% | | |
| Parry, M. L., C. Rosenzweig, A. Iglesias, M. Livermore, and G. Fischer. 2004. “Effects of Climate Change on Global Food Production under SRES Emissions and Socio-Economic Scenarios.” Global Environmental Change. https://doi.org/10.1016/j.gloenvcha.2003.10.008 . | Access Price (food crops) | Price | increase in price | NA | | - | - | 5-35% | - | - | - | | Increase fertiliser and pesticide application, irrigation |
| Fujimori, Shinichiro, Tomoko Hasegawa, Joeri Rogelj, Xuanming Su, Petr Havlik, Volker Krey, Kiyoshi Takahashi, and Keywan Riahi. 2018. “Inclusive Climate Change Mitigation and Food Security Policy under 1.5 °C Climate Goal.” Environmental Research Letters 13 (7): 074033. https://doi.org/10.1088/1748-9326/aad0f7 . | Access Price (food crops) | price | increase in price | NA | | - | - | - | - | - | - | | food policy scenarios (international aid, domestic reallocation, bioenergy tax) |
| Hertel, Thomas W., Marshall B. Burke, and David B. Lobell. 2010. “The Poverty Implications of Climate-Induced Crop Yield Changes by 2030.” Global Environmental Change. https://doi.org/10.1016/j.gloenvcha.2010.07.001 . | Access Price (major staples) | Price | increase in price | NA | | 3.60% | 10-15% | - | - | - | - | | new crop varieties, significant expansion of irrigation Infrastructure |
| UNCCD 2017 | Access (disproportionate impact on low-income consumers, in particular women and girls, due to lack of resources to purchase food) | soil health | negative | NA | | - | - | - | - | - | - | | low (soil health provides key adaptation option, without which lit reviewed by UNCCD points towards low adaptation potential) |
| Vermeulen, Sonja J., Bruce Campbell, and John S. Ingram. 2012. “Climate Change and Food Systems.” SSRN. https://doi.org/10.1146/annurev-environ-020411-130608 . | Access (inability to invest in adaptation and diversification measures to endure price rises) | agricultural yields and earnings, food prices, reliability of delivery, food quality, and, notably, | reduced access to food | NA | | - | - | - | - | - | - | | low |

| | | food safety | | | | | | | | | | | |
|--|---|-------------|------------------------|-------|--|---|---|---|---|---|---|--|-----------------|
| Morris, George Paterson, Stefan Reis, Sheila Anne Beck, Lora Elderkin Fleming, William Neil Adger, Timothy Guy Benton, and Michael Harold Depledge. 2017. "Scoping the Proximal and Distal Dimensions of Climate Change on Health and Wellbeing." Environmental Health: A Global Access Science Source. https://doi.org/10.1186/s12940-017-0329-y . | Access (indirect impacts due to spatial dislocation of consumption from production for many societies) | crop yield | reduced access to food | GGCMs | | - | - | - | - | - | - | strong negative effects of climate change, especially at higher levels of warming and at low latitudes | |
| FAO 2016a | Access (loss of agricultural income due to reduced yields and higher costs of production inputs, such as water, limits ability to buy food) | crop yield | negative | NA | | - | - | - | - | - | - | likely 1.0 and 1.5 | low to moderate |
| Abid, Muhammad, Uwe A. Schneider, and Jürgen Scheffran. 2016. "Adaptation to Climate Change and Its Impacts on Food Productivity and Crop Income: Perspectives of Farmers in Rural Pakistan." Journal of Rural Studies. https://doi.org/10.1016/j.jrurstud.2016.08.005 . | Access (loss of agricultural income due to reduced yields and higher costs of production inputs, such as water, limits ability to buy food) | farm income | negative | NA | | - | - | - | - | - | - | likely 1.0 and 1.5 | low |
| Harvey, Celia A., Zo Lalaina Rakotobe, Nalini S. Rao, Radhika Dave, Hery Razafimahatratra, Rivo Hasinandrianina Rabarijohn, Haingo Rajaofara, and James L. MacKinnon. 2014. "Extreme Vulnerability of Smallholder Farmers to Agricultural Risks and Climate Change in Madagascar." Philosophical Transactions of the Royal Society B: Biological Sciences. https://doi.org/10.1098/rstb.2013.0089 . | Access (loss of agricultural income due to reduced yields and higher costs of production inputs, such as water, limits ability to buy food) | farm income | negative | NA | | - | - | - | - | - | - | likely 1.0 and 1.5 | low |

| | | | | | | | | | | | | | |
|---|---|------------------|--|----|--|---|---|--|---|------|---|--|--|
| Calvin, Katherine, Marshall Wise, Page Kyle, Pralit Patel, Leon Clarke, and Jae Edmonds. 2014. "Trade-Offs of Different Land and Bioenergy Policies on the Path to Achieving Climate Targets." Climatic Change 123 (3-4): 691-704. https://doi.org/10.1007/s10584-013-0897-y . | Access (Price) | Price | increase in price | NA | | - | - | - | - | 320% | - | | |
| Kreidenweis, Ulrich, Florian Humpenöder, Miodrag Stevanović, Benjamin Leon Bodirsky, Elmar Kriegler, Hermann Lotze-Campen, and Alexander Popp. 2016. "Afforestation to Mitigate Climate Change: Impacts on Food Prices under Consideration of Albedo Effects." Environmental Research Letters 11 (8): 085001. https://doi.org/10.1088/1748-9326/11/8/085001 . | Access (Price) | Price | increase in price | NA | | - | - | 60-80% | - | - | - | | Increase investment in R&D, etc |
| Tilman, David, and Michael Clark. 2014. "Global Diets Link Environmental Sustainability and Human Health." Nature. https://doi.org/10.1038/nature13959 . | Access demand | demand | doubling of demands by 2050 | NA | | - | - | - | - | - | - | | |
| Chatzopoulos, Thomas, Ignacio Pérez Domínguez, Matteo Zampieri, and Andrea Toreti. 2019. "Climate Extremes and Agricultural Commodity Markets: A Global Economic Analysis of Regionally Simulated Events." Weather and Climate Extremes. https://doi.org/10.1016/j.wace.2019.100193 . et al., 2019 | Access | Economic impacts | | | negative. Large-scale events will 'very likely' occur more frequently, more intensely, and last longer | key wheat-growing regions display yield reductions from -28% (Australia) to -6% (US and Ukraine). | "Besides Australia, three more regions exceed a reduction of -20%: Canada, Russia, and Kazakhstan." | "persistent large-scale harvest failures may deplete grain stocks and thus render future prices even more responsive." | - | - | | unspecified in the modeling approach based on extreme events, implied 1.5GMS | governments trapped in risk-averse or risk-taking behavior, difficult to achieve and sustain crop stocks to buffer |
| UTILIZATION | | | | | | | | | | | | | |
| Müller, Christoph, Joshua Elliott, and Anders Levermann. 2014. "Food Security: Fertilizing Hidden Hunger." Nature Climate Change. https://doi.org/10.1038/nclimate2290 . | Utilization (decline in nutritional quality resulting from increasing | human migration | negative (heat stress induced long-term migration of people) | NA | | - | - | - | - | - | - | likely between 1.0 and 1.5 due to heat stress peaks | low (unless long term migration is considered an acceptable form of migration) |

| | | | | | | | | | | | | | |
|--|--|--|------------------------------|----|-----------------|---|---|--|---|---|---|---|---|
| | atmospheric CO2) | | | | | | | | | | | | |
| Myers, Samuel S., Antonella Zanobetti, Itai Kloog, Peter Huybers, Andrew D.B. Leakey, Arnold J. Bloom, Eli Carlisle, et al. 2014. "Increasing CO2 Threatens Human Nutrition." Nature. https://doi.org/10.1038/nature13179 . | Utilization (decline in nutritional quality resulting from increasing atmospheric CO2) | zinc and iron | reduced nutrition | NA | 2050 or 550ppm | - | - | - | - | - | - | 550ppm | Low/Moderate. Differences between cultivars of a single crop suggest that breeding for decreased sensitivity to atmospheric CO2 concentration could partly address these new challenges to global health. |
| Smith, M. R., C. D. Golden, and S. S. Myers. 2017. "Potential Rise in Iron Deficiency Due to Future Anthropogenic Carbon Dioxide Emissions." GeoHealth. https://doi.org/10.1002/2016gh000018 . | Utilization (decline in nutritional quality resulting from increasing atmospheric CO2) | iron | negative (iron deficiency) | NA | | - | - | 550 ppm | - | - | - | likely between 1.0 and 1.5 due to heat stress peaks | low to moderate |
| Myers, Samuel S., K. Ryan Wessells, Itai Kloog, Antonella Zanobetti, and Joel Schwartz. 2015. "Effect of Increased Concentrations of Atmospheric Carbon Dioxide on the Global Threat of Zinc Deficiency: A Modelling Study." The Lancet Global Health. https://doi.org/10.1016/S2214-109X(15)00093-5 . | Utilization (decline in nutritional quality resulting from increasing atmospheric CO2) | zinc deficiency under different CO2 concentrations | negative (zinc deficiency) | NA | 2050 | - | - | The total number of people estimated to be placed at new risk of zinc deficiency by 2050 was 138 million (95% CI 120-156). | - | - | - | | moderate |
| Moretti, Antonio, Michelangelo Pascale, and Antonio F. Logrieco. 2019. "Mycotoxin Risks under a Climate Change Scenario in Europe." Trends in Food Science and Technology. https://doi.org/10.1016/j.tifs.2018.03.008 . | Utilization (higher post-harvest losses due to mycotoxins) | crops after harvest | reduced availability of food | NA | current to 2050 | - | - | - | - | - | - | possibly between 1.0 and 1.5 | low to moderate |
| Fels-Klerx, H.J. Van der, C. Liu, and P. Battilani. 2016. "Modelling Climate Change Impacts on Mycotoxin | Utilization (negative impact on food | crops after harvest | reduced utilization of food | NA | | - | - | - | - | - | - | likely between 1.0 and 1.5 | not yet clear |

| | | | | | | | | | | | | | | |
|---|---|--|---------------------------------|----|---------------------|--|--|---|---|---|---|--|--|--------------------------|
| Contamination.” World Mycotoxin Journal. https://doi.org/10.3920/wmj2016.2066 . | safety due to effect of increased temperatures on microorganisms, including increased mycotoxins in food and feed) | | | | | | | | | | | | | |
| Tirado, Maria Cristina, and Janice Meerman. 2012. “Climate Change and Food and Nutrition Security.” In The Impact of Climate Change and Bioenergy on Nutrition. https://doi.org/10.1007/978-94-007-0110-6-4 . | Utilization (negative impact on food safety due to effect of increased temperatures on microorganisms, including increased mycotoxins in food and feed) | | reduced utilization of food | NA | to mid century | - | - | - | - | - | - | | | moderate |
| Aberman, Noora Lisa, and Cristina Tirado. 2014. “Impacts of Climate Change on Food Utilization.” In Global Environmental Change. https://doi.org/10.1007/978-94-007-5784-4_124 . | Utilization (negative impact on nutrition resulting from reduced water quantity and quality used to prepare food) | food availability, utilization, access | negative | NA | 2020-end of century | - | - | - | - | - | - | likely between 1.0 and 1.5 | | low (water availability) |
| Thompson, Brian, and Marc J. Cohen. 2012. The Impact of Climate Change and Bioenergy on Nutrition. The Impact of Climate Change and Bioenergy on Nutrition. https://doi.org/10.1007/978-94-007-0110-6 . | Utilization (negative impact on nutrition resulting from reduced water quantity and quality used to prepare food) | nutrition, distribution of food | negative | NA | | - | - | - | - | - | - | | | low |
| Special Report on Global Warming of 1.5°C Summary for Policymakers, 2018 | Utilization (nutrition) | nutrients | Decrease in nutritional content | NA | | at 0.87, yellow - associated impacts are both detectable | associated impacts are both detectable and attributable to | indicates closer to severe and widespread | - | - | - | Limiting global warming to 1.5°C compared to 2°C | | |

| | | | | | | ble and attributable to climate change with at least medium confidence. | climate change with at least medium confidence. | impacts. | | | | would result in a lower global reduction in nutritional quality |
|---|-----------------------|-----------|---|----|--|---|---|----------|---|---|---|--|
| Bahrami, Helale, Luit J. De Kok, Roger Armstrong, Glenn J. Fitzgerald, Maryse Bourgault, Samuel Henty, Michael Tausz, and Sabine Tausz-Posch. 2017. "The Proportion of Nitrate in Leaf Nitrogen, but Not Changes in Root Growth, Are Associated with Decreased Grain Protein in Wheat under Elevated [CO2]." Journal of Plant Physiology. https://doi.org/10.1016/j.jplph.2017.05.011 . | Utilization Nutrients | nutrients | above ground biomass production and yield will typically increase by 17–20% while concentrations of nutrients such as N will decrease by 9–15% in plant tissues. Here they found - The 12% loss in grain protein under e[CO2] | NA | | - | - | - | - | - | - | Grain yield per plant was greater under e[CO2]. Irrigation treatment significantly enhanced grain yield by 128%. Grain protein concentration (%) decreased by 12% in e[CO2] grown wheat compared to a[CO2]. Grain protein concentration (%) was 15% higher in rain-fed than well-watered treatments but did not differ between the two wheat cultivars. Continuing favourable water supply conditions for photosynthesis during grain filling can prolong carbohydrate delivery to grains and thereby increase yield but depress grain protein, which is consistent with greater |

| | | | | | | | | | | | | | |
|---|----------------------------------|---------------------|---|----|-----------|--------------------------|----------------------------------|---|---|---|---|--|---|
| | | | | | | | | | | | | | grain yield and lower grain protein concentrations in wellwatered compared to rain-fed crops in our study |
| Medek, Danielle E., Joel Schwartz, and Samuel S. Myers. 2017. "Estimated Effects of Future Atmospheric Co2 concentrations on Protein Intake and the Risk of Protein Deficiency by Country and Region." Environmental Health Perspectives. https://doi.org/10.1289/EHP41 . | Utilization nutrition | protein content | Decrease Under eCO2, rice, wheat, barley, and potato protein contents decreased by 7.6%, 7.8%, 14.1%, and 6.4%, respectively. | NA | | - | - | - | - | - | - | | |
| Smith, M. R., C. D. Golden, and S. S. Myers. 2017. "Potential Rise in Iron Deficiency Due to Future Anthropogenic Carbon Dioxide Emissions." GeoHealth. https://doi.org/10.1002/2016gh000018 . | Utilization nutrition | nutrients | CO2 concentrations of 550 ppm can lead to 3–11% decreases of zinc and iron concentrations in cereal grains and legumes and 5–10% reductions in the concentration of phosphorus, potassium, calcium, sulfur, magnesium, iron, zinc, copper, and manganese across a wide range of crops under more extreme conditions of 690 ppmCO2 | NA | | - | - | - | - | - | - | | |
| Puma, Michael J., Satyajit Bose, So Young Chon, and Benjamin I. Cook. 2015. "Assessing the Evolving Fragility of | Utilization (disruptions to food | crops after harvest | reduced utilization of food | NA | 1992-2009 | moderate risk at present | increased connectivity and flows | - | - | - | - | | low |

| | | | | | | | | | | | | | | |
|---|--|-------------|--|----|--|--|--|---|---|---|---|--------------------|--|--|
| the Global Food System.” Environmental Research Letters. https://doi.org/10.1088/1748-9326/10/2/024007 . | storage and transportation networks) | | | | | | within global trade networks suggest that the global food system is vulnerable to systemic disruptions, especially considering tendency for exporting countries to switch to non-exporting states during times of food scarcity in the global markets. | | | | | | | |
| Wellesley, Laura, Felix Preston, Johanna Lehne, and Rob Bailey. 2017. “Chokepoints in Global Food Trade: Assessing the Risk.” Research in Transportation Business and Management. https://doi.org/10.1016/j.rtbm.2017.07.007 . | Utilization (disruptions to food storage and transportation networks) | food prices | reduced utilization of food | NA | | - | - | - | - | - | - | likely 1.0 and 1.5 | moderate | |
| STABILITY | | | | | | | | | | | | | | |
| Schmidhuber, J., and F. N. Tubiello. 2007. “Global Food Security under Climate Change.” Proceedings of the National Academy of Sciences. https://doi.org/10.1073/pnas.0701976104 ., 2007 | Stability | | High Fluctuation (price, supply, yields) | NA | | negative. increased fluctuations in crop yields and local food supplies and higher risks of landslides and erosion damage, they can adversely affect the stability of food supplies and thus | In semiarid areas, droughts can dramatically reduce crop yields and livestock numbers and productivity (most in sub-Saharan Africa and parts of South Asia) poorest regions with the highest level of | - | - | - | - | | Food import, freer trade, investment (storage, irrigation, transport, communication) | |

| | | | | | | | | | | | | | |
|--|--|----------------|--------------------------------|----|------|---------------|--|---|---|---|---|--|----------|
| | | | | | | food security | chronic undernourishment will also be exposed to the highest degree of instability in food production | | | | | | |
| Zheng, Jingyun, Lingbo Xiao, Xiuqi Fang, Zhixin Hao, Quansheng Ge, and Beibei Li. 2014. "How Climate Change Impacted the Collapse of the Ming Dynasty." Climatic Change. https://doi.org/10.1007/s10584-014-1244-7 . | Stability (civil disturbance, social tension) | social tension | disruption food supply | NA | | - | 1. Extreme events will severely disrupt the food supply 2. Extreme events will escalate popular unrest, rebellions and wars 2. Extreme events will increase expenditure to 60 -70% | - | - | - | - | | |
| Diffenbaugh, Noah S., Thomas W. Hertel, Martin Scherer, and Monika Verma. 2012. "Response of Corn Markets to Climate Volatility under Alternative Energy Futures." Nature Climate Change. https://doi.org/10.1038/nclimate1491 . | Stability (impacts on world market export prices that carry through to domestic consumer prices due to climate shocks) | price of corn | negative | NA | | - | - | - | - | - | - | | low |
| Verma, Monika, Thomas Hertel, and Noah Diffenbaugh. 2014. "Market-Oriented Ethanol and Corn-Trade Policies Can Reduce Climate-Induced US Corn Price Volatility." Environmental Research Letters. https://doi.org/10.1088/1748-9326/9/6/064028 . | Stability (impacts on world market export prices that carry through to domestic consumer prices due to climate shocks) | price of corn | likely negative | NA | | - | - | - | - | - | - | | low |
| Willenbockel, Dirk. 2012. Extreme Weather Events and Crop Price Spikes in a Changing | Stability (impacts on world market | food price | negative (potential food price | NA | 2030 | - | 1. Extreme events, such as | - | - | - | - | | moderate |

| | | | | | | | | | | | | | | |
|---|---|-----------------------|--|----|---------|-----------------|--|---|---|---|---|--|--|---|
| Climate, Illustrative Global Simulation Scenarios. Oxfam Research Reports. | export prices that carry through to domestic consumer prices due to climate shocks) | | impacts of a number of extreme weather event scenarios in 2030 for each of the main exporting regions for rice, maize and wheat) | | | | flooding , can wipe out economic infrastructure; 2. Agricultural infrastructure will be affected 3. weather-related yield shocks occurred will occur 4. Global crop production will drop | | | | | | | |
| Salmon, J.Meghan, Mark A. Friedl, Steve Frolking, Dominik Wisser, and Ellen M. Douglas. 2015. "Global Rain-Fed, Irrigated, and Paddy Croplands: A New High Resolution Map Derived from Remote Sensing, Crop Inventories and Climate Data." International Journal of Applied Earth Observation and Geoinformation. https://doi.org/10.1016/j.jag.2015.01.014 . | stability (political and economic) | rainfall, temperature | disruption food supply, price fluctuation, decrease in production | NA | | - | - | - | - | - | - | | | agricultural intensification, changes in land use practices |
| Medina-Elizalde, Martín, and Eelco J. Rohling. 2012. "Collapse of Classic Maya Civilization Related to Modest Reduction in Precipitation." Science. https://doi.org/10.1126/science.1216629 . | stability (political and economic) | rainfall | Low yields | NA | | - | - | - | - | - | - | | | |
| Challinor, Andy J., W. Neil Adger, Tim G. Benton, Declan Conway, Manoj Joshi, and Dave Frame. 2018. "Transmission of Climate Risks across Sectors and Borders." Philosophical Transactions of the Royal Society A: Mathematical, Physical and Engineering Sciences. https://doi.org/10.1098/rsta.2017.0301 . | Stability (widespread crop failure contributing to migration and conflict) | crop failure | negative | NA | | - | - | - | - | - | - | | | moderate |
| Hendrix, Cullen S. 2018. "Searching for Climate–conflict Links." Nature Climate Change. https://doi.org/10.1038/s41558-018-0083-3 . | Stability (widespread crop failure contributing to migration and conflict) | crop failure | negative | NA | current | - | - | - | - | - | - | | | moderate |
| Kelley, Colin, Shahrzad Mohtadi, Mark Cane, | Stability (widespr | crop failure | negative | NA | current | negative.severe | "Multiyear | - | - | - | - | | | low to medium. |

| | | | | | | | | | | | | | | |
|--|--|----------------------|---|----|---------|--|---|-------------------------------|------------------------------|---|---|--------------------------------------|---|---|
| Richard Seager, and Yochanan Kushnir. 2017. "Commentary on the Syria Case: Climate as a Contributing Factor." Political Geography. https://doi.org/10.1016/j.polgeo.2017.06.013 . | lead crop failure contributing to migration and conflict) | | | | | drought 2006/2007 caused northeastern "breadbasket" region to collapse (zero or near-zero production, livestock herds lost). | drought episodes in the late 1950s, 1980s, and 1990s, the total population of Syria (Fig. 1D) grew from 4 million in the 1950s to 22 million in recent years; (ii) decline ground water supply (iii) drought occurred shortly after the 1990s drought | | | | | | | |
| Kelley, Colin P., Shahrzad Mohtadi, Mark A. Cane, Richard Seager, and Yochanan Kushnir. 2015. "Climate Change in the Fertile Crescent and Implications of the Recent Syrian Drought." Proceedings of the National Academy of Sciences 112 (11): 3241–46. https://doi.org/10.1073/pnas.1421533112 . | Stability (widespread crop failure contributing to migration and conflict) | crop failure | negative, low yields and price increase | NA | current | - | 1. Extreme events will lead to unprecedented rise in food prices 2. Extreme events will obliterate livestock | - | - | - | - | | | low |
| Schmidhuber, J., and F. N. Tubiello. 2007. "Global Food Security under Climate Change." Proceedings of the National Academy of Sciences. https://doi.org/10.1073/pnas.0701976104 . | Stability (production, supply chain, extreme events) | extreme events | Fluctuation (yield and supply), Reduction (labour, productivity), Increase (disease burden) | NA | | - | 1. droughts can dramatically reduce crop yields and livestock productivity 2. exposed to the highest degree of instability in food production | - | - | - | - | | | Food imports, Freer trade, Investment (storage, irrigation, transport, communication) |
| Chatzopoulos, Thomas, Ignacio Pérez Domínguez, Matteo Zampieri, and Andrea Toreti. 2019. "Climate Extremes and | stability (variability in supply, price) | yield, market, price | Fluctuation (yield, market and price) | NA | | negative. climate extremes collide | key wheat-growing regions display yield | Besides Australia, three more | The transmission of domestic | | - | unspecified in the modeling approach | buffer stock schemes for stabilizing supply | |

| | | | | | | | | | | | | | |
|--|--------------------------|-----------------------------|--|-----------|------------------|--|--|--|---|--|--|--|---|
| <p>Agricultural Commodity Markets: A Global Economic Analysis of Regionally Simulated Events.” Weather and Climate Extremes. https://doi.org/10.1016/j.wace.2019.100193, et al., 2019</p> | | | | | | <p>with major drivers (population growth, dietary shifts, environmental degradation, and trade interdependence</p> | <p>reductions –28% (Australia) to –6% (US and Ukraine).</p> | <p>regions exceeded a reduction of 20%: Canada, Russia, and Kazakhstan. The highest absolute drops, corresponding to –0.9 t/ha and –0.7 t/ha, were found in Canada and Russia.</p> | <p>prices to global markets is visible in most scenarios with large shocks in key exporters and importers being responsible for the most pronounced effects</p> | | | <p>h based on extreme events, implied 1.5GM ST. “Economic simulation models typically operate under the assumption of ‘normal’ growing conditions, contain no explicit parameterization of climatic anomalies on the supply side, and confound multifarious sources of yield fluctuation in harvest-failure scenarios”</p> | <p>and prices of major staple commodities in food-insecure regions may mitigate some of the induced price volatility but are generally difficult to achieve and sustain in practice</p> |
| <p>Bellemare, Marc F. “Rising Food Prices, Food Price Volatility, and Social Unrest.” American Journal of Agricultural Economics, 2015, doi:10.1093/ajae/aau038.</p> | <p>Stability (trade)</p> | <p>trade, supply, price</p> | <p>negative, trade in situations where global grain production is reduced does not distribute world food stocks / inadequate and counter to modeling results (in reality producing countries protect domestic grain reserves; prices</p> | <p>NA</p> | <p>2007-2010</p> | <p>negative.</p> | <p>1990-2011 food price increases led to increases in social unrest, food price volatility has not been associated with increases in social unrest</p> | | | | | | <p>medium in SSP1-like world</p> |

| | | | | | | | | | | | | | | |
|--|--|----------------------|--|----|-----------|---|---|---|---|---|---|----------|--------|--|
| | | | spike upwards in times of reduced yields but do not fall as much in times of normal or increased yields) | | | | | | | | | | | |
| Zampieri, M., A. Ceglar, F. Dentener, and A. Toreti. 2017. "Wheat Yield Loss Attributable to Heat Waves, Drought and Water Excess at the Global, National and Subnational Scales." Environmental Research Letters. https://doi.org/10.1088/1748-9326/aa723b . | stability(variability in supply, price) | yield, market, price | Fluctuation (yield, market and price) | NA | | negative. | - | - | - | - | - | | | |
| Donati, Michele, et al. "The Impact of Investors in Agricultural Commodity Derivative Markets." Outlook on Agriculture, 2016, doi:10.5367/oa.2016.0233. | Stability (trade) | trade, supply, price | negative, trade in situations where global grain production is reduced does not distribute world food stocks / inadequate and counter to modeling results (in reality producing countries protect domestic grain reserves; prices spike upwards in times of reduced yields but do not fall as much in times of normal or increased yields) | | 2007-2010 | negative. | open trade helps improve access to food at lower prices, combined with observations in other articles about impact of market speculation (US) combined with export restraints (Russia, Ukraine, India, Vietnam) in 2007-2011 drought periods. | - | - | - | - | | | |
| Gilbert, C. L., and C. W. Morgan. "Food Price Volatility." Philosophical Transactions of the Royal Society B: Biological Sciences, 2010, | Stability (trade) | trade, supply, price | negative, trade in situations where global grain production is | | 2007-2010 | negative. not yet clear if trend in food price volatility | "World dollar prices of major agricultural food commod | - | - | - | - | moderate | global | |

| | | | | | | | | | | | | | |
|--|--------------------------|-----------------------------|--|--|------------------|---|---|----------|----------|----------|----------|--|---|
| <p>doi:10.1098/rstb.2010.0139.</p> | | | <p>reduced does not distribute world food stocks / inadequate and counter to modeling results (in reality producing countries protect domestic grain reserves; prices spike upwards in times of reduced yields but do not fall as much in times of normal or increased yields)</p> | | | <p>y is permanent</p> | <p>ities rose dramatically from late 2006 through to mid-2008. Prices collapsed dramatically in the second half of 2008 with the onset of the financial crisis. periods of high volatility have been relatively short and interspaced with longer periods of market tranquility. It would therefore be wrong simply to extrapolate recent and current high volatility levels into the future. However, it remains valid to ask whether part of the volatility rise may be permanent."</p> | | | | | | |
| <p>Gilbert, Christopher L. "How to Understand High Food Prices." <i>Journal of Agricultural Economics</i>, 2010, doi:10.1111/j.1477-9552.2010.00248.x.</p> | <p>Stability (trade)</p> | <p>trade, supply, price</p> | <p>negative, trade in situations where global grain production is reduced does not</p> | | <p>2007-2010</p> | <p>negative. not yet clear if trend in food price volatility is permanent</p> | <p>index-based investment in agricultural futures markets is seen as the</p> | <p>-</p> | <p>-</p> | <p>-</p> | <p>-</p> | | <p>moderate depending on exposure to market speculation</p> |

| | | | | | | | | | | | | |
|--|----------------------|----------------------------|---|--|---------------|--|---|---|---|---|--|---|
| | | | distribut e world food stocks / inadequa te and counter to modelin g results (in reality producin g countries protect domestic grain reserves; prices spike upwards in times of reduced yields but do not fall as much in times of normal or increase d yields) | | ent | major channel through which macroec onomic and monetar y factors generate d the 2007– 2008 food price rise | | | | | | |
| Headey, Derek. "Rethinking the Global Food Crisis: The Role of Trade Shocks." Food Policy, 2011, doi:10.1016/j.foodpol.20 10.10.003. | Stability (trade) | trade, supply, price | negative, trade in situation s where global grain producti on is reduced does not distribut e world food stocks / inadequa te and counter to modelin g results (in reality producin g countries protect domestic grain reserves; prices spike upwards in times of reduced yields but do not fall as much in times of normal or increase d yields) | | negativ e. | when food prices peaked in June of 2008, they soared well above the new equilibri um price. observat ions that internati onal rice prices surged in response to export restricti ons by India and Vietnam suggeste d that trade- related factors could be an importa nt basis for oversho oting, especiall y given the very tangible | "In all cases except soybea ns, we find that large surges in export volum es preced ed the price surges. The presen ce of these large deman d surges, togethe r with back- of-the- envelo pe estimat es of their price impact s, sugges ts that trade events played a much larger and | - | - | - | | monthly data from Thailand (the largest exporter of rice), and the United States (the largest exporter of wheat and maize and the third largest exporter of soybeans). |

| | | | | | | | | | | | | | |
|--|------------------------------|----------------------|--|-----------|--|--|---|---|---|---|---|---|--|
| | | | | | | link between export volumes and export prices | more pervasive role than previously though t." | | | | | | |
| Marchand, Philippe, et al. "Reserves and Trade Jointly Determine Exposure to Food Supply Shocks." Environmental Research Letters, 2016, doi:10.1088/1748-9326/11/9/095009. | Stability (trade) | trade, supply, price | negative, trade in situations where global grain production is reduced does not distribute world food stocks / inadequate and counter to modeling results (in reality producing countries protect domestic grain reserves; prices spike upwards in times of reduced yields but do not fall as much in times of normal or increased yields) | 2007-2010 | negative, without coordinated and effective international and domestic risk management of food stocks. | supply shocks driven not only by the intensification of trade, but as importantly by changes in the distribution of reserves . trade dependency may accentuate the risk of food shortages from foreign production shocks | increased number and volume of trade links (relative to production), decrease and a more even distribution of global reserves (still relative to production). ->distribution of reserves matters more than their aggregate quantity in terms of conferring resilience to shocks . | Possibility of multiple supply side shocks across different regions of the world (multi-breadbasket failure) | Compounded risk: Trade greater reliance on imports increases the risk of critical food supply losses following a foreign shock, notably in the case of several Central American and Caribbean countries that import grains from the United States " | - | | Medium. Trade dependency has substantially increased in the last few decades and more than doubled since the mid-1980s (Porkka et al 2013, D'Odorico et al 2014) likely as a result of liberalization and the associated removal of subsidies and trade protections in developing countries (e.g., Shafaeddin 2005)." | |
| Sternberg, Troy. "Chinese Drought, Bread and the Arab Spring." Applied Geography, 2012, doi:10.1016/j.apgeog.2012.02.004. | Stability (trade, political) | trade, supply, price | negative, trade in situations where global grain production is reduced does not distribute world food stocks / inadequate and counter to modeling results (in | 2007-2010 | "Chinese drought contributed to a doubling of global wheat prices. The drought affected the price of bread in Egypt which influenced | - | - | - | - | - | - | - | Depends on food reserves, trade policy (risk management) and if multi-breadbasket failure is present |

| | | | | | | | | | | | | | | |
|----------------------------------|------------------------|---|---|--|----------------------------|---|---|---|---|---|---|---|---|--------|
| | | | reality producing countries protect domestic grain reserves; prices spike upwards in times of reduced yields but do not fall as much in times of normal or increased yields) | | | political protest. The process exemplifies the potential global consequences of climate hazards today." | | | | | | | | |
| Permafrost degradation | | | | | | | | | | | | | | |
| Chadburn, S. E. et al., 2017 NCC | permafrost degradation | Permafrost area change (million km ²) | increased loss of permafrost, leading to radical changes in high-latitude hydrology and biogeochemical cycling. Estimated sensitivity of permafrost area loss to global mean warming at stabilization of 4.0 +/- 1.1 million km ² °C ⁻¹ . | CMIP5, multiple RCPs | 1850-2300 | Indirectly | 13 | 9 | 6 | 4 | 2 | - | - | Global |
| Burke, E. J. et al., 2018 ERL | permafrost degradation | Increased land carbon emissions at stabilization Gt C /yr | Additional emissions between 225 and 345 GtC (10th to 90th percentile) from permafrost thaw under 2 °C stabilised warming .60–100 GtC less in a 1.5 | JULES-IMOGEN intermediate complexity climate model | 1.5° and 2°C stabilization | - | 1.5: 0.08 to 0.16 Gt C yr ⁻¹ (10th to 90th percentile) | 0.09 to 0.19 GtC yr ⁻¹ (10th to 90th percentile) | - | - | - | - | - | Global |

| | | | | | | | | | | | | | | |
|----------------------------|------------------------|---------------------------------------|--|--------|-----------|---|--|---|--|--|---|---|---|---|
| | | | °C world. | | | | | | | | | | | |
| Jorgenson & Osterkamp 2005 | permafrost degradation | Water erosion | Increased water erosion | Review | - | - | - | - | - | - | - | - | - | Global |
| Gauthier et al., 2015 | permafrost degradation | Tree mortality | Permafrost thawing in dry continental Siberia may trigger widespread drought-induced mortality in dark coniferous forests and larch forests that cover 20% of the global boreal forest | Review | - | - | - | - | - | - | - | - | - | Fennoscandia, Siberia and the northern reaches of North America |
| FAO 2012 | permafrost degradation | Damage to forest hydrological regimes | Permafrost thawing will reinforce the greenhouse effect and induce irreversible damage to forest hydrological regimes, especially across regions receiving little rainfall. | Review | 2012-2030 | - | - | - | - | Carbon release by 2100 could be several times that of current tropical deforestation | - | - | - | Siberia |
| Price et al., 2013 | permafrost degradation | Permafrost thaw | Increases in nearsurface permafrost temperatures during 2007–2009 are up to 2 °C warmer compared to 2-3 decades, and there is a concurrent trend in its degradation | Review | 1995-2100 | - | Permafrost is now warming at almost all sites across the North American permafrost zones, except for site where the permafrost is already close to | - | Rapid degradation and disappearance over extensive areas within next 50–100 years (Camil 2005; Smith et al. 2005). Accelerated degradation | 16%–35% of Canadian permafrost area in 2000 may be lost by 2100 (Zhang et al., 2008a; 2008b) | - | - | - | Canada |

| | | | | | | | | | | | | | |
|----------------------------|------------------------|--|--|------------------------------------|-------------|---|--|---|--|---|---|---|---------------|
| | | | on and disappearance. Overall transient responses of permafrost to warming are likely to be nonlinear. | | | | 0 °C and vertical ground temperature profiles are isothermal, indicating ongoing phase changes (Smith et al. 2010) | | ation by 2050 likely in several regions. | | | | |
| Hjort et al., 2018 NatComm | permafrost degradation | Proportion of all residential, transportation, and industrial infrastructure in areas of nearsurface permafrost thaw (a) and high hazard (b) in the pan-Arctic permafrost area (%) | Arctic infrastructure at risk from degrading permafrost by mid-century | Infrastructure hazard computations | 2041 – 2060 | - | - | 4 million people, 70% of current infrastructure | - | - | - | - | Global |
| Fire | | | | | | | | | | | | | |
| Bajocco et al., 2010 | fire | Area burned | Multidirectional relationships between climate, land degradation and fire may be amplified under future land use change and climate scenarios (Bajocco et al. 2010). | - | 1990-2000 | - | - | - | - | - | - | - | Mediterranean |
| Marlon et al., 2016 | fire | Biomass burning | Increase in charcoal influx (i.e. biomass burning) during the industrial period (probably not related to climate | Paleoclimate reconstruction | 22ka-2000 | - | - | - | - | - | - | - | Global |

| | | | | | | | | | | | | | |
|-----------------------------|------|-------------|---|---------------------|-----------|---------------------------|---|---|---|---|---|---------------------|--|
| | | | but human activities) | | | | | | | | | | |
| Giglio et al., 2016 | fire | Area burned | Trends in land area burnt have varied regionally | Recent observations | 1995-2011 | Regionally varying trends | - | - | - | - | - | - | Northern Hemisphere Africa has experienced a fire decrease of 1.7 Mha yr-1 (-1.4% yr-1) since 2000, while Southern Hemisphere Africa saw an increase of 2.3 Mha yr-1 (+1.8% yr-1) during the same period. Southeast Asia witnessed a small increase of 0.2 Mha yr-1 (+2.5% yr-1) since 1997, while Australia experienced a sharp decrease of about 5.5 Mha yr-1 (-10.7% yr-1) during 2001-11, followed by an upsurge in 2011 that exceeded the annual area burned in the previous 14 years |
| Andela et al., 2017 Science | fire | Area burned | A recent analysis using the Global Fire Emissions Database v.4 that includes small fires concluded that the net reduction in land | Remote sensing | 1998-2015 | Global decline | - | - | - | - | - | high in the tropics | Global |

| | | | | | | | | | | | | | |
|-------------------------------|------|--------------------|---|-----------------------|-----------|---|---|---|---|---|---|--|----------------------------------|
| | | | area burnt globally during 1998–2015 was -24.3±8.8% (-1.35 ± 0.49% yr ⁻¹). However, from the point of fire emissions it is important to consider the land cover types which have experienced changes in area burned; in this instance, most of the declines have come from grasslands, savannas and other non-forest land cover types (Andela et al. 2017). | | | | | | | | | | |
| Abatzoglou and Williams, 2016 | fire | Forest area burned | Significant recent increases in forest area burned (with higher fuel consumption per unit area) recorded in western and boreal North America. | Detection/attribution | 1979-2015 | plus 100% cumulative forest fire area, CC accounted for 55% of increase in fuel aridity | - | - | - | - | - | moderate (rise in forest fires despite increasing adaptation measures) | western and boreal north America |
| Ansmann et al., 2018 | fire | Forest area burned | Clear link between the western Canadian fires and | Aerosoles, case study | 2017-2017 | - | - | - | - | - | - | - | western and boreal north America |

| | | | | | | | | | | | | | |
|---------------------------|------|---|---|--|-----------|-----------------------------------|-----------|-----------|-----------|------------|---------|-------------------------------|---|
| | | | aerosol loading over Europe. | | | | | | | | | | |
| Pechony and Shindell 2010 | fire | Fire activity (% rel to pre-industrial) | Temperature increase and precipitation decline may become the major driver of fire regimes under future climates as evapotranspiration increases and soil moisture decreases. | Driving forces, A2, A1B, B1; single GCM, AR4-era | 800-2100 | - | plus0-10% | plus0-10% | plus5-10% | plus10-35% | plus15% | low under high warming levels | "Although temperatures rise throughout the country, it becomes more humid and rainy in the East and drier in the West (Fig. 4B). Consequently, in the eastern United States fire activity declines, while rising considerably in the western United States (Fig. 4A). In both cases increasing population densities and land-cover changes (Fig. 4C) generally reduce fire activity." |
| Aldersley et al., 2011 | fire | Fire regimes | Temperature increase and precipitation decline may become the major driver of fire regimes under future climates as evapotranspiration increases and soil moisture decreases. | Random forest on data sets | 2000-2000 | - | - | - | - | - | - | - | Global |
| Fernandes et al., 2017 | fire | Fire regimes | Temperature increase and precipitation | Logistic regression | 1995-2015 | Yes, for Indonesia during moderat | - | - | - | - | - | - | Indonesia |

| | | | | | | | | | | | | | |
|--------------------|------|----------------------------|---|------------------|-----------|----------------|------------|---|---|---|---|---|---|
| | | | tion decline may become the major driver of fire regimes under future climates as evapotranspiration increases and soil moisture decreases. | | | e to wet years | | | | | | | |
| Liu et al., 2010 | fire | Probability of fire | The risk of wildfires in future could be expected to change, increasing significantly in North America, South America, central Asia, southern Europe, southern Africa, and Australia | KBDI on GCM data | 2070-2100 | - | - | - | - | - | - | - | North America, South America, central Asia, southern Europe, southern Africa, and Australia |
| Jolly et al., 2015 | fire | Fire weather season length | Fire weather season has already increased by 18.7% globally between 1979 and 2013, with statistically significant increases across 25.3% but decreases only across 10.7% of Earth's land surface covered with vegetation; even sharper changes have | Weather analysis | 1979-2013 | Yes, global | plus 18.7% | - | - | - | - | - | Global |

| | | | | | | | | | | | | | |
|-------------------------|------|--|---|-----------------------------|-----------|-------------|------------|-----------|-----------|--------|---------|---|--------|
| | | | been observed during the second half of this period. | | | | | | | | | | |
| Jolly et al., 2015 | fire | Area experiencing long weather fire season | Global area experiencing long weather fire season has increased by 3.1% per annum or 108.1% during 1979–2013. | Weather analysis | 1979-2013 | Yes, global | plus108.1% | - | - | - | - | - | Global |
| Huang et al., 2014 | fire | Fire frequencies | Fire frequencies by 2050 are projected to increase by ~27% globally, relative to the 2000 levels, with changes in future fire meteorology playing the most important role in enhancing the future global wildfires, followed by land cover changes, lightning activities and land use, while changes in population density exhibits the opposite effects. | A1B | 2000-2050 | - | - | - | 19% | - | - | - | Global |
| Knorr et al., 2016a NCC | fire | Area burned | Climate is only one driver of a complex set of | SIMFIRE+LPJGUESS RCP4.5/8.5 | 1971-2100 | - | no change | no change | no change | plus5% | plus10% | - | Global |

| | | | | | | | | | | | | | |
|-------------------------|------|--------------------|--|------------------------------|-----------|---|-----|---|---------|---|---------|---|--------|
| | | | <p>environmental, ecological and human factors in influencing fire (Bowman et al. 2011). Interplay leads to complex projections of future burnt area and fire emissions (Knorr et al. 2016b,a), yet human exposure to wildland fires is projected to increase because of population expansion into areas already under high risk of fires.</p> | | | | | | | | | | |
| Knorr et al., 2016a NCC | fire | Exposure (#people) | <p>Climate is only one driver of a complex set of environmental, ecological and human factors in influencing fire (Bowman et al. 2011). Interplay leads to complex projections of future burnt area and fire emissions (Knorr et al. 2016b,a), yet human exposure</p> | SIMFIRE +LPIGUESS RCP4.5/8.5 | 1971-2100 | - | 413 | - | 497-646 | - | 527-716 | - | Global |

| | | | | | | | | | | | | | |
|------------------------|------|------------------------------------|--|------------------------------|-----------|---|------|---|---|---|---|---|----------------------------|
| | | | to wildland fires is projected to increase because of population expansion into areas already under high risk of fires. | | | | | | | | | | |
| Knorr et al., 2016b BG | fire | Greenhouse gas emissions from fire | Climate is only one driver of a complex set of environmental, ecological and human factors in influencing fire (Bowman et al. 2011). Interplay leads to complex projections of future burnt area and fire emissions (Knorr et al. 2016b,a), yet human exposure to wildland fires is projected to increase because of population expansion into areas already under high risk of fires. | SIMFIRE +LPJGUESS RCP4.5/8.5 | 1971-2100 | - | -15% | - | - | - | - | - | Global |
| Flannigan et al., 2009 | fire | Area burned, fire season length | General increase in area burned and fire occurrence but a lot of spatial variability, with | Review | pre-2100 | - | - | - | - | - | - | - | Review of regional studies |

| | | | | | | | | | | | | | |
|---------------------------------|------|--|--|---|-----------|------------------------------|------|--------|--------|---|---|---|---|
| | | | some areas of no change or even decreases in area burned and occurrence. Fire seasons are lengthening for temperate and boreal regions and trend will continue in a warmer world. Future trends of fire severity and intensity are difficult to determine owing to the complex and non-linear interactions between weather, vegetation and people. | | | | | | | | | | |
| Abatzoglou et al., 2019 | fire | Multimodel median proportion of burnable terrestrial surfaces for which emergence occurs (%) | Anthropogenic increases in extreme Fire Weather Index days emerge for an increasingly large fraction of burnable land area under higher global temperatures. | Fire Weather Index on 17 CMIP5 climate models | 1861-2099 | Yes, on 22% of burnable land | 0-3% | 15-30% | 30-50% | - | - | - | Global (pronounced effects in Mediterranean and Amazon) |
| Westerling et al., 2006 Science | fire | Wildfire frequency and duration | Higher large-wildfire frequency, longer wildfire durations, and longer wildfire | Fire reports | 1970-2003 | Yes, for Western US | - | - | - | - | - | - | Western US |

| | | | | | | | | | | | | | |
|------------------------------------|---------------------|---|---|------------------------------------|-----------------------|----|----------------------------|------------|-------------|---|----------|---|---|
| | | | seasons. | | | | | | | | | | |
| Yang et al., 2014 JGR | fire | Area burned | Global decline in recent burned area (1.28 × 104km2 yr1), driven significant decline in tropics and extratropics caused by human factors. warming and droughts are expected to increase wildfire activity towards the future. | DLEM-Fire | 1901-2007 | - | - | - | - | - | - | - | Global |
| Turco, M. et al., 2018 | fire | Area burned | Increase in burned area scales with warming levels. Substantial benefits from limiting warming to well below 2 °C. | SM and NSM under RCP2.6 and RCP8.5 | 1981-2100 | - | - | plus50-75% | plus75-175% | - | - | - | Mediterranean |
| Flannigan et al., 2005 | fire | Area burned | Increase burned area under enhanced CO2 scenarios | 2xCO2, 3xCO2 (cfr SRES A2) | 1975-1995; 2050; 2100 | - | - | - | plus78% | - | plus143% | - | Canada |
| Coastal degradation | | | | | | | | | | | | | |
| Mentaschi et al., 2018 | coastal degradation | Coastal erosion area (km2) | Substantial global-scale increases in coastal erosion in recent decades. | Remote sensing | 1984 – 2015 | No | 28,000 km2 eroded globally | - | - | - | - | - | Global |
| Neumann, B., et al., 2015 Plos One | coastal degradation | Number of people exposed to a 1-in-100 year flood event incoastal regions (million) | Increase population exposure to 1-in-100 year storm surge. Strongest changes | Population projections | 2000-2060 | No | 625 | 879-949 | 1053-1388 | - | - | - | Coastal regions are also characterized by high population density, particularly in Asia (Banglade |

| | | | | | | | | | | | | | |
|--------------------------------|---------------------|--------------------------------------|--|---|-------------|------------|--------|---|---|---|-------------------|---|---|
| | | | in exposure in Egypt and sub-Saharan countries in Western and Eastern Africa. | | | | | | | | | | sh, China, India, Indonesia, Vietnam) whereas the highest population increase of coastal regions is projected in Africa (East Africa, Egypt, and West Africa) |
| Nicholls et al. 2011 | coastal degradation | Number of people displaced (million) | Increases in coastal erosion. | DIVA model framework | 2000-2100 | No | - | - | - | - | 72-187 (0.9-2.4%) | high: most of the threatened population could be protected. | Global |
| Cazenave and Cozannet 2014 | coastal degradation | - | Increases in coastal erosion. | Review, mostly qualitative | 2000-2100 | No | - | - | - | - | - | - | Global (with Southeast Asia concentrating many locations highly vulnerable to relative sea level rise) |
| Rahmstorf 2010 | coastal degradation | - | Increases in coastal erosion. | Commentary | 2000-2100 | Yes | - | - | - | - | - | - | Global |
| Meeder and Parkinson 2018 | coastal degradation | Coastal erosion | Increases in coastal erosion. | Sedimentary record | 1900-2000 | - | - | - | - | - | - | - | Everglades, USA |
| Shearman et al. 2013 | coastal degradation | Coastal erosion | Net contraction in mangrove area | Land cover classification | 1980s-2000s | Indirectly | -0.28% | - | - | - | - | - | Asia-Pacific Region |
| McInnes et al. 2011 | coastal degradation | Coastal erosion | CMIP3 wind speed exhibit low skill over land areas. | CMIP3 evaluation wind speed, SRES | 1981-2100 | - | - | - | - | - | - | - | Global |
| Mori et al. 2010 | coastal degradation | Coastal erosion | Wave heights increase in future climates across mid-latitudes and the Antarctic Ocean. | GCM combined with a wave model under SRES | 1979-2099 | - | - | - | - | - | - | - | Global (rise in wave height in midlatitudes and southern ocean, decrease in tropics) |
| Savard et al., 2009 | coastal degradation | Coastal erosion | Increases in coastal erosion | Stakeholder discussions | 2005-2007 | - | - | - | - | - | - | - | Canada |
| Tamarin-Brodsky and Kaspi 2017 | coastal degradation | Tropical cyclones | Poleward shift in | Storm tracking | 1980-2099 | - | - | - | - | - | - | - | Midlatitudes |

| | | | | | | | | | | | | | |
|----------------------|---------------------|--|---|--|----------------------|--------------------------------------|-------|----------|-------------|---|---|---|---|
| | ion | | the genesis latitude and increased latitudinal displacement of tropical cyclones under global warming . | algorithm to CMIP5 | | | | | | | | | |
| Ruggiero 2013 | coastal degradation | Total water level | Increases in wave height (and period), increasing the probability of coastal flooding/erosion more than sea level rise alone. | Simple total water level model | 1965-2010 | - | - | - | - | - | - | - | U.S. Pacific Northwest |
| Elliott et al., 2014 | coastal degradation | Nexus | Nexus of climate change and increasing concentration of people . | Review, mostly qualitative ly | - | - | - | - | - | - | - | - | Global |
| Knutson et al., 2010 | coastal degradation | Tropical cyclone intensity | Increase d intensity and frequency of high-intensity hurricanes with higher warming levels. | Review | 1950-2100 | Yes globally , regional ly difficult | - | - | - | - | - | - | Tropical cyclone regions |
| Bender et al., 2010 | coastal degradation | Atlantic hurricane category 4 frequency | Increase d intensity and frequency of high-intensity hurricanes with higher warming levels. | CMIP3 downscaling with hurricane model; SRES A1B | 2001-2020; 2081-2100 | - | - | - | plus75 -81% | - | - | - | Atlantic (with the largest increase projected over the Western Atlantic, north of 20°N) |
| Vecchi et al., 2008 | coastal degradation | Hurricane Power Dissipation Index Anomaly (10 ¹¹ m ³ s ⁻²) | Increase d intensity and frequency of high-intensity hurricanes with higher warming | Statistical regression SST PDI applied to CMIP | 1950-2100 | - | plus1 | -1 to +4 | -1 to +6 | - | - | - | Atlantic |

| | | | | | | | | | | | | | |
|-------------------------|---------------------|---|--|------------------------------|----------------------|---|------------|-------------|---|---|---|---|--------------------------|
| | | | levels. | | | | | | | | | | |
| Bhatia et al., 2018 | coastal degradation | Tropical cyclone category 4 frequency (# TCs) | Frequency, intensity, and intensification distribution of TCs all shift to higher values during the twenty-first century. | RCP4.5, single GCM | 2016-2035; 2081-2100 | - | plus26-67% | plus27-133% | | - | - | - | Tropical cyclone regions |
| Bhatia et al., 2018 | coastal degradation | Tropical cyclone category 5 frequency (# TCs) | Frequency, intensity, and intensification distribution of TCs all shift to higher values during the twenty-first century. | RCP4.5, single GCM | 2016-2035; 2081-2100 | - | plus46-50% | plus85-200% | | - | - | - | Tropical cyclone regions |
| Tu et al., 2018 | coastal degradation | Tropical cyclones | Regime shift in the destructive potential of tropical cyclones around 1998, with regional regulation by the ElNiño/Southern Oscillation and the Pacific Decadal Oscillation. | PDI on observations | 1979-2016 | No | - | - | - | - | - | - | Western North Pacific |
| Sharmila and Walsh 2018 | coastal degradation | Tropical cyclones paths | Tropical cyclones paths shift poleward | Reanalysis | 1980-2014 | Indirectly: hadley cell expansion has been linked to climate change | - | - | - | - | - | - | Tropical cyclone regions |
| Kossin 2018 | coastal degradation | Tropical cyclones translation speed | Over the last seven decades, the speed at which tropical | Best-track data from IBTrACS | 1949-2016 | Indirectly: trend analysis | - | - | - | - | - | - | Tropical cyclone regions |

| | | | | | | | | | | | | | |
|----------------------|---------------------|--|--|--|------------|--|-------------|---|--------------|---|---|---|-------------|
| | | | cyclones move has decreased significantly as expected from theory, exacerbating the damage on local communities from increasing rainfall amounts | | | | | | | | | | |
| Luke et al., 2016 | coastal degradation | Forest composition | The heterogeneity of land degradation at coasts that are affected by tropical cyclones can be further enhanced by the interaction of its components (for example, rainfall, wind speed, and direction) with topographic and biological factors (for example, species susceptibility) | Case studies of TC impacts on vegetation | 2004-2007 | - | - | - | - | - | - | - | West Indies |
| Emmanuel 2005 Nature | coastal degradation | Tropical cyclone Power Dissipation Index | Potential destructiveness of hurricanes has increased markedly since the mid-1970s due to both longer storm lifetimes and greater storm intensities. | 'best track' tropical data sets | 1930-2010 | Indirectly: consistency with increase in SST | - | - | - | - | - | - | Global |
| Emmanuel 2017 PNAS | coastal degradation | Tropical cyclone | Increase in | downscaling of large | 1981-2000; | - | x6 increase | - | x18 increase | - | - | - | Texas |

| | ion | precipitation | intense precipitation associated with tropical cyclones | numbers of tropical cyclones from three climate reanalyses and six climate models | 2081-2100 | | in probability since late 20th century | | e in probability since late 20th century | | | | |
|--------------------------------|---------------------|--|--|--|-------------|---|--|---------------------|--|---|---|--|--------------------------|
| Wehner, M. F. et al., 2018 ESD | coastal degradation | Tropical cyclone counts of category 4/5 | Increase in frequency and intensity of most intense tropical cyclones under 1.5°C and 2°C warming levels. | single GCM, HAPPI protocol | HAPPI | - | at 1.5°C: plus 2.1/ plus 1.2 | plus 1.4 / plus 1.2 | - | - | - | - | Tropical cyclone regions |
| Hanson et al., 2011 CC | coastal degradation | People exposed to 1-in-100-year coastal flooding (# people) | Enhanced exposure to extreme coastal flooding, with total population exposure possibly increasing threefold by 2070. | Global rise of 0.5 m above current levels by 2070, +10% increase in extreme water levels | 2005; 2070s | - | 38.5 M people (0.6%) | 150 M people | - | - | - | high! "This research shows the high potential benefits from risk-reduction planning and policies at the city scale to address the issues raised by the possible growth in exposure." (paper) | Global |
| Hanson et al., 2011 CC | coastal degradation | Assets exposed to 1-in-100-year coastal flooding (% global GDP of that period) | Enhanced exposure to extreme coastal flooding, with total population exposure possibly increasing threefold by 2070. | Global rise of 0.5 m above current levels by 2070, +10% increase in extreme water levels | 2005; 2070s | - | 5% | 9% | - | - | - | high! "This research shows the high potential benefits from risk-reduction planning and policies at the city scale to address the issues raised by the possible | Global |

| | | | | | | | | | | | | | | |
|-------------------------------|---------------------|---|---|------------------------------|-----------------|---|---------|--------------|---|--------------|---|---|------------------------------|--------|
| | | | | | | | | | | | | | growth in exposure." (paper) | |
| Vousdoukas et al., 2016 CDD | coastal degradation | Extreme storm surge levels | The anticipated increase in relative sea level rise can be further enforced by an increase in extreme storm surge levels. | RCP4.5 + 8.5, 8 CMIP5 models | 1970-2100 | - | - | - | - | - | - | - | present and needed | Europe |
| Vousdoukas et al., 2017 EF | coastal degradation | Extreme sea level change compared to present-day | 100-year extreme sea level along Europe's coastline is on average projected to increase by 57/81 cm for RCP4.5/8.5. | RCP4.5 + 8.5, 6 CMIP5 models | 1980-2014; 2100 | - | - | plus57 cm | - | plus81cm | - | - | - | Europe |
| Vousdoukas et al., 2017 EF | coastal degradation | Extreme sea level return period affecting 5 Million Europeans | 100-year extreme sea level along Europe's coastline is on average projected to increase by 57(81) cm for RCP4.5(8.5). | RCP4.5 + 8.5, 6 CMIP5 models | 1980-2014; 2100 | - | 100year | 3 year | - | 1 year | - | - | - | Europe |
| Vousdoukas et al., 2018 NComm | coastal degradation | Extreme sea level change compared to present-day | By 2050, extreme sea level rise would annually expose a large part of the tropics to the present-day 100-year event. Unprecedented flood risk levels by the end of the century unless | RCP4.5 + 8.5, 6 CMIP5 models | 1980-2014; 2100 | - | - | plus34-76 cm | - | plus58-172cm | - | - | - | Global |

| | | | | | | | | | | | | | |
|--------------------------------|---------------------|---|--|---|------------|---|---------------------------|------------------------|-----------------------------|-------|---|---|---|
| | | | timely adaptation measures are taken. | | | | | | | | | | |
| Rasmussen, D. J. et al., 2018 | coastal degradation | Human population exposure under 2150 local SLR projections (millions) | Increased permafrost melt, increased coastal erosion | 1.5K, 2.0K, 2.5K stabilisation scenarios | 2010; 2150 | - | 1.5: 56.05 (32.54–112.97) | 61.84 (32.89 – 138.63) | 2.5: 62.27 (34.08 – 126.95) | - | - | - | Global |
| Moftakhari et al., 2017 PNAS | coastal degradation | Coastal flooding | Compound flooding from river flow and coastal water level enhances risk derived from univariate assessments. | RCP4.5 + 8.5 | 2030; 2050 | - | - | - | - | - | - | - | Global |
| van den Hurk et al., 2015 ERL | coastal degradation | Coastal flooding | Compound flooding from river flow and coastal water level enhances risk derived from univariate assessments. | 800 sim years with an RCM | 2012-2012 | - | - | - | - | - | - | - | The Netherlands |
| Zscheischler et al., 2018 NCC | coastal degradation | Coastal flooding | Interaction between multiple climate drivers and/or hazards play a major role in coastal extremes. | Review | - | - | - | - | - | - | - | - | USA |
| Jevrejeva, S. et al., 2018 ERL | coastal degradation | Coastal flooding | Rising global annual flood costs with future warming. | 1.5K, 2.0K, stabilisation + RCP8.5 in CMIP5 | 2100 | - | 1.5°C: 1 | 1.2 | - | 14-27 | - | "Adaptation could potentially reduce sea level induced flood costs by a factor of 10" (paper) | Global, "Upper middle income countries are projected to experience the largest increase in annual flood |

| | | | | | | | | | | | | | |
|------------------------------|---------------------|--|--|---|-----------------------|---|----------------------|---------------|-----------|---|---------------|--|--|
| | | | | | | | | | | | | | costs (up to 8% GDP) with a large proportion attributed to China. High income countries have lower projected flood costs, in part due to their high present-day protection standards." (paper) |
| Brown, S. et al., 2018 REC | coastal degradation | Decadal average of land inundated by flooding (km ²) | Increased soil erosion, increased soil salinity, subsidizing land with future warming. | 1.5, 2.0 and 3.0 stabilization scenarios from SRES A1B, with Delta Dynamic Integrated Emulator Model | 1986-2005; 2050; 2100 | - | 1.5°C: 1000-1500 | 1500-1700 | 2000-2500 | - | - | "With slow rates of sea-level rise, adaptation remains possible, but further support is required" (paper) | Ganges-Brahmaputra-Meghna and other vulnerable deltas |
| Nicholls, R. J. et al., 2018 | coastal degradation | Expected people flooded (millions yr ⁻¹) | Increase in coastal inundation and number of people exposed under future warming levels. | 1.5K, 2.0K, stabilization scenarios + RCP8.5 in CMIP5; Warming Acidification and Sea Level Projector Earth systems model, large ensembles | 1986-2300 | - | 1.5°C: 150 (100-230) | 170 (120-270) | - | - | 400 (220-700) | "adaptation remains essential in densely populated and economically important coastal areas under climate stabilization. Given the multiple adaptation steps that this will require, an adaptation pathway approach has merits for coastal areas." | Global |

| | | | | | | | | | | | | | |
|------------------------------|---------------------|--|---|--|---|---|-------|---|---|---------------|---|---|--|
| | | | | | | | | | | | | (paper) | |
| Mentaschi et al., 2017 GRL | coastal degradation | Extreme wave energy flux in 100yr return level | More extreme wave activity in the southern hemisphere towards the end of the century. | Spectral wave model Wavewatch III forced by 6 CMIP5 models under RCP8.5 | 1980-2010; 2070-2100 | - | - | - | - | up to plus30% | - | - | Southern hemisphere |
| Villarini et al., 2014 BAMS | coastal degradation | Coastal flooding | Flooding from tropical cyclones affects large areas of the United States. | Discharge measurements | 1981-2011 | - | - | - | - | - | - | - | Eastern US |
| Woodruff et al., 2013 Nature | coastal degradation | Coastal flooding | Increase in future extreme flood elevations. | Review of global and regional studies | 1981-2100 | - | - | - | - | - | - | - | Global |
| Brecht et al., 2012 JED | coastal degradation | Coastal flooding | Strong inequalities in the risk from future disasters. | Implications of tropical storm intensification for 31 developing countries and 393 of their coastal cities with populations greater than 100,000 | 2000-2100 | - | - | - | - | - | - | - | Selected cities across the world |
| Hallegatte et al 2013 | coastal degradation | Flood losses (Billion US\$ yr-1) | Increasing global flood future warming | Quantification of present and future flood losses in the 136 largest coastal cities. | 2005; 2050 (20 and 40 cm sea level rise; assume 2°C but no info in paper) | - | 6 | 1000 without adaptation, 60-63 with adaptation keeping constant flood probability | - | - | - | huge challenge: "To maintain present flood risk, adaptation will need to reduce flood probabilities below present values" (paper) | Global |
| Jongman et al., 2012 GEC | coastal degradation | People and value of assets in flood-prone regions (Trillion US\$ in 1/00 coastal flood hazard areas) | Increase in people and asset exposure in 1-in-100-year coastal flood hazard areas. | Population density and GDP per capita estimate; land-use estimate | 2010; 2050 | - | 27-46 | 80-158 | - | - | - | - | Global (largest population exposure increase in Asia (absolute) and Sub-Saharan+North Africa (relative)) |

| | | | | | | | | | | | | | |
|-------------------------------|---------------------|---|--|---|---------------------|----------------------|---------|---|---|---|---|---|----------------|
| Muis et al., 2018 EF | coastal degradation | Coastal flooding | Significant correlations across the Pacific between ENSO and extreme sea levels. | Tides and storm surge reanalysis | 1979-2014 | No | - | - | - | - | - | - | Global |
| Reed et al., PNAS | coastal degradation | Return period of 1/500yr pre-industrial flood height (yr) | Mean flood heights increased by ~1.24 m from ~A.D.850 to present. | Proxy sea level records and downscaled CMIP5 | 850-1800; 1970-2005 | Yes | 24 year | - | - | - | - | - | New York |
| Wahl et al., NCC | coastal degradation | Return period of 1/100yr pre-industrial flood height (yr) | Increase in the number of coastal compound events over the past century . | Statistical analyses | 1900-2012 | Yes | 42 year | - | - | - | - | - | USA & New York |
| Vegetation degradation | | | | | | | | | | | | | |
| Allen et al., 2010 | vegetation changes | Tree mortality | Increases in tree mortality | Global assessment of recent tree mortality attributed to drought and heat stress. "Although episodic mortality occurs in the absence of climate change, studies compiled here suggest that at least some of the world's forested ecosystems already may be responding to climate change and raise concern that forests may become increasingly vulnerable to higher background tree mortality | 1970-2008 | Yes but not formally | - | - | - | - | - | - | quasi-Global |

| | | | | | | | | | | | | | | |
|-----------------------|--------------------|---|--|---|-----------|----------------------|--|---|---|---|---|---|---|---|
| | | | | rates and die-off in response to future warming and drought, even in environments that are not normally considered water-limited" (paper) | | | | | | | | | | |
| Trumbore et al., 2015 | vegetation changes | Forest health | Intensification of stresses on forests | Review | - | - | - | - | - | - | - | - | - | - |
| Hember et al., 2017 | vegetation changes | Net ecosystem biomass production (NEBP) | A 90% increase in NEBP driven by environmental changes. | Observations at 10,307 plots across southern ecozones of Canada | 1501-2012 | Yes but not formally | rise in wet climates, decline in dry climates | - | - | - | - | - | - | Canada |
| Midgley and Bond 2015 | vegetation changes | Vegetation structure | Climate, atmospheric CO2 and disturbance changes are able to shift vegetation between states. | Review | - | - | - | - | - | - | - | - | - | Africa |
| Norby et al., 2010 | vegetation changes | Net Primary Productivity (NPP, kg dry matter m ⁻² yr ⁻¹) | Increasing N limitation, expected from development and exacerbated by elevated CO2. | FACE: CO2 vs N | 1998-2008 | - | reduction in NPP difference between ambient and elevated CO2 experiments | - | - | - | - | - | - | High latitudes |
| Gauthier et al., 2015 | vegetation changes | Boreal forest shift to woodland/shrubland biome | Increase in drought-induced mortality, changes in climate and related disturbances may overwhelm the resilience of species and ecosystems, possibly leading to important | Review | - | - | climate zones shift faster than adaptation capacity | - | - | - | - | - | - | Fennoscandia, Siberia and the northern reaches of North America |

| | | | | | | | | | | | | | |
|-----------------------|--------------------|----------------------------|---|------------------|-----------|---|--|---|---|---|------------------------------------|---|--|
| | | | t biome-level changes. | | | | | | | | | | |
| FAO 2012 | vegetation changes | Boreal forest productivity | Enhanced dieback and timber quality decrease despite increase in forest productivity. | Review | 2012-2030 | - | "Higher forest mortality is already being observed in practically all areas of the boreal belt." | - | - | - | mass destruction of forest stands. | "The state of knowledge regarding adaptive potential and the regional vulnerability of forests to climate change is insufficient" (paper) | Siberia (highest risks for Southern regions and forest steppe) |
| Price et al., 2013 | vegetation changes | Boreal forest productivity | Where precipitation is generally nonlimiting, warming coupled with increasing atmospheric carbon dioxide may stimulate higher forest productivity. Increase in large wildfires. Risk of endemic forest insect pests population outbreaks in response to relatively small temperature increases. | Review | 1995-2100 | - | - | - | - | - | - | - | Canada |
| Girardin et al., 2016 | vegetation changes | Boreal forest productivity | Tree growth dependence on soil moisture in boreal Canada since the mid-20th century. Projections of | Dendrochronology | 1950-2015 | drought and heat control boreal tree growth | no change | - | - | - | - | - | North America |

| | | | | | | | | | | | | | |
|-----------------------|--------------------|---------------------------------|---|---|-----------|---------------------------------------|---|---|---|---|-------------|---|---|
| | | | future drying pose risk to forests especially in moisture-limited regimes. | | | | | | | | | | |
| Beck et al., 2011 | vegetation changes | Boreal forest productivity | Growth increases at the boreal-tundra ecotones in contrast with drought-induced productivity declines throughout interior Alaska. Initiating biome shift. | Dendrochronology and remote sensing | 1982-2010 | drought-induced productivity declines | - | - | - | - | - | - | North America |
| Lewis et al., 2004 | vegetation changes | Tropical forest health | Widespread changes observed in mature tropical forests. | Review | 1900-2001 | - | - | - | - | - | - | - | Global |
| Bonan et al., 2008 | vegetation changes | Forest health | Forests under large pressure from global change. | Review | - | - | - | - | - | - | - | - | Global |
| Miles et al., 2004 | vegetation changes | Species becoming non-viable (%) | Little change in the realized distributions of most species due to delays in population responses. | HADCM2 GSa1 1%CO2 (old ref) | 1990-2095 | - | - | - | - | - | 43% by 2095 | - | Amazonia (highest risks over lowland and montane forests of Western Amazonia) |
| Anderegg et al., 2012 | vegetation changes | Tree mortality | Increased tree mortality | Review | - | - | - | - | - | - | - | - | Global |
| Sturrock et al., 2011 | vegetation changes | Tree mortality | Increased tree mortality | Review "We Review knowledge of relationships between climate variables and several forest diseases, as well as current evidence | - | - | - | - | - | - | - | "Regardless of these uncertainties, impacts of climate change on forest health must be mitigated. This will require | Global |

| | | | | | | | | | | | | |
|--|--|--|--|--|--|--|--|--|--|--|--|--|
| | | | | of how climate, host and pathogen interactions are responding or might respond to climate change." (paper) | | | | | | | | proactive thinking and a modified suite of forest management approaches, because status quo management strategies will not protect forest values in a changing climate. Climate change is already disrupting practices and policies for managing commercial and non-commercial forests, such as forest classification systems, projections of growth and yield and subsequent models of supply for timber and other forest products, plans and projections for managing habitat for differen |
|--|--|--|--|--|--|--|--|--|--|--|--|--|

| | | | | | | | | | | | | | |
|--------------------|-----------------------|-------------------|---------------------------------|--|---------------|---|---|---|--|---|---|---|------------------|
| | | | | | | | | | | | | t species of animals , and cycling of carbon, nutrient s and water (Graham et al., 1990)." (paper) | |
| Bentz et al., 2010 | vegetation changes | Tree mortality | Increase d tree mortality | Population models forced with CRCM climate projections under A2 | 1961- 2100 | - | - | - | e.g. Spruce beetle: "In the period 2001– 2030 and again from 2071 to 2100, we would expect substa ntial increas es in spruce forest area with high probab ility of spruce beetle offspri ng produc ed annual ly rather than semian nually (figure 1b, 1c, 1e, 1f). By the end of the centur y, the change in temper atures across the boreal forests of central Canada may cause markedly | - | - | - | North America |

| | | | | | | | | | | | | | | | | | | |
|-----------------------|--------------------|----------------|---|------------------|-----------|--|---|---|---|---|---|---|---|---|--|--|--|------------------------|
| | | | | | | | | | | | | | | higher probability of spruce beetle outbreak potential, based on developmental timing alone. A model for predicting the cold tolerance of this insect is not available" (paper) | | | | |
| McDowell et al., 2011 | vegetation changes | Tree mortality | Increased tree mortality | Synthetic theory | 1850-2100 | - | - | - | - | - | - | - | - | | | | | Global |
| Lindner et al., 2010 | vegetation changes | Tree mortality | positive effects on forest growth and wood production from increasing atmospheric CO2 content and warmer temperatures especially in northern and western Europe. Increasing drought and disturbance (e.g. fire) risks will cause adverse effects, outweighing positive trends in southern and eastern Europe. | Review | 2000-2100 | Some changes already detected (e.g. in Pyrenees) | - | - | - | - | - | - | - | | | | | Europe |
| Mokria et al. 2015 | vegetation changes | Tree mortality | Decreasing trend in tree | Dendrochronology | 2006-2013 | - | - | - | - | - | - | - | - | | | | | Northern Ethiopia, dry |

| | | | | | | | | | | | | | |
|-------------------------------|--------------------|--|--|--|-----------|---|---|---|---|---|---|---|--------------------|
| | | | mortality with increasing elevation. | | | | | | | | | | afromontane forest |
| Shanahan et al., 2016 | vegetation changes | Abrupt woodland - grassland shifts | Interactions between climate, CO2 and fire can make tropical ecosystems more resilient to change, but systems are dynamically unstable and potentially susceptible to abrupt shifts between woodland and grassland dominated states in the future. | 28,000-year integrated record of vegetation, climate and fire from West Africa | 15-28Ka | - | - | - | - | - | - | - | West Africa |
| Ferry Slik et al., 2002 | vegetation changes | Tree mortality | Reduction in number of trees and tree species per surface area directly after disturbance (fire). | Forest plot monitoring | 1970-2002 | - | - | - | - | - | - | - | Indonesia |
| Dale et al., 2001 | vegetation changes | Tree mortality | Altered frequency, intensity, duration and timing of fires, droughts, introduced species and other disturbances can affect forests. | Review | - | - | - | - | - | - | - | - | Global |
| Schlesinger and Jasechko 2014 | vegetation changes | ratio of transpiration over evapotranspiration (%) | Changes in transpiration due to rising CO2 | Review | - | - | - | - | - | - | - | - | Global |

| | | | | | | | | | | | | | |
|------------------------|---------------------|---|---|--|------------|--------------------------------------|---|-----|------------------------------|---|---|---|------------------------|
| | | | concentrations, land use changes, shifting ecozones and climate warming . | | | | | | | | | | |
| Loucks et al., 2010 CC | coastal degradation | Number of breeding tiger species | Tiger habitat loss under future climate change. High agreement that the joint effect of climate change and land degradation will be very negative for the area. | Sea level rise scenarios of 0, 12, 28cm (assumed 1,2,3K) | 2000-2090 | - | 115 | 105 | 5 | - | - | - | Sundarbhan, Bangladesh |
| Payo et al., 2016 CC | coastal degradation | Mangrove area loss (km2) | Increasing mangrove area losses by 2100 relative to 2000 due to sea level rise. | Sea level rise scenarios of 0.46, 0.75 and 1.48m | 2000; 2100 | - | - | - | 81-1391k m ² lost | - | - | - | Sundarbhan, Bangladesh |
| Song et al., 2018 | vegetation changes | Land change | 60% of all recent land changes are associated with direct human activities whereas 40% with indirect drivers such as climate change. | Remote Sensing | 1982-2016 | - | 40% of land change from indirect drivers such as climate change | - | - | - | - | - | Global |
| Mc Kee et al. 2004 GEB | vegetation changes | Salt marsh dieback (ha) | Vegetation dieback and soil degradation. | Areal and ground surveys | 2000-2001 | - | More than 100,000 ha affected, with 43,000 ha severely damaged | - | - | - | - | - | USA |
| soil erosion | | | | | | | | | | | | | |
| Li and Fang, 2016 | soil erosion | Soil erosion rates (t ha ⁻¹ yr ⁻¹) | more often than not studies project an | Review | 1990-2100 | Indirectly: close links demonstrated | 0-73.04 | - | - | - | - | - | Global |

| | | | | | | | | | | | | | |
|-----------------------------|--------------|---|--|---|----------------------|--|----------|---------|----------|----------|---|---|------------------|
| | | | increase in erosion rates (+1.2 to +1600%, 49 out of 205 studies project more than 50% increase) | | | regionally, no formal D&A | | | | | | | |
| Serpa et al., 2015 | soil erosion | Sediment export change in humid/dry catchment (%) | Decrease in streamflow (2071-2100) | SWAT + ECHAM SRES A1B and B1 | 1971-2000; 2071-2100 | - | - | -22/+5% | -29/+22% | - | - | - | Mediterranean |
| Neupane and Kumar, 2015 | soil erosion | Change in river flow | Dominant effect of LULCC | SWAT under SRES B1, A1B, A2 | 1987-2001; 2091-2100 | - | - | - | - | - | - | - | Big Sioux River |
| Mullan et al., 2012 | soil erosion | Change in soil erosion | Erosion rates without land management changes would decrease by 2020s, 2050s and 2100s, dominant effect of land management | WEPP under SRES | 2020s; 2050s; 2080s | - | - | - | - | - | - | - | Northern Ireland |
| Bond-Lamberty et al., 2018 | soil erosion | Soil organic matter (SOM) | Soil carbon decline | Global soil respiration data base | 1990-2014 | - | - | - | - | - | - | - | Global |
| Bellmay et al., 2005 Nature | soil erosion | Soil property changes | Soil carbon decline | National soil inventory of England and Wales | 1978-2003 | Indirectly: relation between rate of carbon loss and carbon content irrespective of land use, suggesting a link to climate change. | -0.6%/yr | - | - | - | - | - | UK |
| Ramankutty et al., 2002 | soil erosion | Suitability for agriculture (%) | Increase in suitability for agriculture in northern high latitudes | IS92a 'business as usual' "calibrating the satellite-based IGBP-DIS 1-km land-cover | 1992; 2070-2099 | - | - | - | - | plus 16% | - | - | Global |

| | | | | | | | | | | | | | | |
|-------------------------------------|--------------|--|--|---|----------------------|---|---|---|---|----------|---|---|---|---------------|
| | | | decrease in tropics | classification dataset (Loveland et al., 2000) against a worldwide collection of agricultural census data." (paper) | | | | | | | | | | |
| Zabel et al., 2014 | soil erosion | Suitability for agriculture (million km ²) | Increase in suitability for agriculture in northern high latitudes, decrease in tropics | ECHAM5 SRES A1B | 1980-2010; 2071-2100 | - | - | - | - | plus 5.6 | - | - | - | Global |
| Burt et al., 2016b | soil erosion | Extreme precipitation indices | Soil erosion may increase in a warmer, wetter world, yet land management is first-order control. | Commentary | 1900-2016 | - | - | - | - | - | - | - | - | India |
| Capolongo et al., 2008 | soil erosion | Climate erosivity | Influence on soil erosion in Mediterranean | Simplified rainfall erosivity model | 1951-2000 | - | - | - | - | - | - | - | - | Mediterranean |
| Barring et al. 2003 Catena | soil erosion | Wind erosion | No clear trend in wind erosion. | Review | 1901-2000 | - | - | - | - | - | - | - | - | Sweden |
| Munson et al., 2011 PNAS | soil erosion | Wind erosion | Enhanced wind erosion. | Wind erosion model | 1989-2008 | - | - | - | - | - | - | - | - | USA |
| Allen & Breshears 1998 - PNAS | soil erosion | Water erosion | Increase in water erosion. | Observations | 1950-1990 | - | - | - | - | - | - | - | - | USA |
| Shakesby 2011 Earth Science Reviews | soil erosion | Water erosion | Water erosion after wildfire not notably distinct in Mediterranean, likely due to land use effects | Review | - | - | - | - | - | - | - | - | - | Mediterranean |
| Pruski and Nearing 2002 | soil erosion | Water erosion | Complex interactions between several factors that affect erosion. | HadCM3 | 1990-2099 | - | - | - | - | - | - | - | - | USA |

| | | | | | | | | | | | | | |
|---|--------------|---|---|---|------------------|---|--|---|-------------|---|---|---|-------------|
| Knorr et al., 2005 Nature | soil erosion | Soil Organic Carbon (SOC) turnover time | Soil carbon decline | Three-pool model, theoretical study | - | - | - | - | - | - | - | - | Global |
| Bond-Lamberty & Thompson 2010 Nature | soil erosion | Soil respiration change (PgC yr ⁻¹) | Enhanced soil respiration. | Database of worldwide soil respiration observations | 1961-2008 | - | plus0.1 Pg C/yr | - | - | - | - | - | Global |
| Jiang et al., 2014 | soil erosion | Soil erosion rates (t ha ⁻¹ yr ⁻¹) | No significant change in soil erosion during one decade | Revised Universal Soil Loss Equation (RUSLE) | 2000; 2006; 2012 | - | - | - | - | - | - | - | Mount Elgon |
| Vanmaercke et al. 2011 (Science of the Total Environment) | soil erosion | Sediment yield | High sediment yield indicates desertification. | Review | - | - | - | - | - | - | - | - | Europe |
| Vanmaercke et al. 2016 (Earth-Science Reviews) | soil erosion | Volumetric gully headcut retreat rate change (%) | Increase in headcut retreat rates | Gully headcut retreat sensitivity to climate | - | - | gully erosion already forms an important problem in many regions | - | plus27-300% | - | - | - | Global |
| de Vente et al. 2013 ESR | soil erosion | Soil erosion and sediment yield | Importance of spatial and temporal scales when considering erosion processes. | Review | - | - | - | - | - | - | - | - | Global |
| Broeckx et al., 2018 ESR | soil erosion | Landslide susceptibility | precipitation not a significant driver of landslide susceptibility, but is significant in non-arid climates | Review | - | - | - | - | - | - | - | - | Africa |
| Gariano and Guzetti 2016 ESR | soil erosion | Landslide susceptibility | Increase in the number of people exposed to landslide risk in regions with future enhanced frequency | Review | - | - | - | - | - | - | - | - | Global |

| | | | | | | | | | | | | | |
|-----------------------------------|----------------|--------------|---|---------------|-------------------------------------|--|---|-----------|-----------|---|---|---|----------------------|
| | | | y and intensity of severe rainfall events. | | | | | | | | | | |
| Water scarcity in drylands | | | | | | | | | | | | | |
| IPCC AR5 | water scarcity | drought | | observations | historical | high confidence in observed trends in some regions of the world, including drought increases in the Mediterranean and West Africa and drought decreases in central North America and northwest Australia | | | | | | | |
| Hoegh-Guldberg et al., 2018 | water scarcity | drought | | observations | historical | medium confidence that greenhouse forcing has contributed to increased drying in the Mediterranean region (including southern Europe, northern Africa and the Near East) | | | | | | | |
| Greve et al., GRL, 2015 | water scarcity | P-ET (mm) | generally a decrease in P-ET in dryland regions but not statistically significant | RCP8.5 | 2080 – 2099 compared to 1980 – 1999 | - | - | - | - | - | - | - | global |
| Byers et al., ERL, 2018 | water scarcity | water stress | increased water | time sampling | 2050 | - | - | 391 (11%) | 418 (12%) | - | - | - | Drylands particularl |

| | | | | | | | | | | | | | |
|--|--------------------------------|---|---|--|--------------------------------------|---|---|---|------------|---|--|---|--|
| | | index (population exposed and vulnerable in drylands, in millions and in percentage of drylands population) | stress with temperature | approach using a combination of RCPs | | | | | | | | | y impacted, including southwestern North America, southeastern Brazil, northern Africa, the Mediterranean, the Middle East, and western, southern and eastern Asia |
| Hanasaki, N., et al. 2013, Hydrol. Earth Syst. Sci., 17, 2393–2413, doi:10.5194/hess-17-2393-2013. | water scarcity | percentage of population under severely water-stressed conditions based on Cumulative Abstraction to Demand ratio CAD≤0.5 | increase with time and RCP | RCP2.6, 4.5, 8.5 | (2071 – 2100 compared to 1971 – 2000 | - | - | 3.6% - 12% | 6.2% - 16% | - | 12.3% - 22.4% | - | global |
| Huang, J. et al. 2017 (NCC) Drylands face potential threat under 2C global warming target (CarbonBrief) | impact of temperature increase | temperature | higher temperature increase in drylands compared to rest of the world | | | - | - | 44% more warming over drylands than humid lands | - | - | - | - | drylands/global |
| Zeng and Yoon, GRL, 2009 | increase desert area | expansion of desert area (i.e. LAI less than 1) | increase in desert area | A1B | 2099 compared to 1901 | - | - | - | - | - | 2.5 million km2 (10% increase)/ with vegetation - albedo feedback: +8.5 million km2 (34% increase) | - | drylands/global |
| Liu, W. et al. 2018 (ESD) Global drought and severe drought-Affected populations in 1.5 and 2C warmer worlds | water scarcity | increase in population exposed to severe drought | increase in exposed population globally | time sampling approach at 1.5 and 2 degree | | - | - | 194.5± 276.5 M | - | - | - | - | global |
| Naumann, G. et al. (2018) Global Changes in Drought Conditions | water scarcity | drought magnitude | increase in drought | time sampling approach | | - | - | Doubling of drought | - | - | - | - | global |

| | | | | | | | | | | | | | |
|-----------------------------------|----------------------------|--|---|---------------------|--|---|---|--|---|---|---|---|--------|
| Under Different Levels of Warming | | | magnitu de | at 1.5 and 2 degree | | | | t magnit ude for 30% of global landma s | | | | | |
| Schewe et al., 2014 PNAS | water scarcity | river runoff as a proxi for water resources | increase in populati on confront ed to water scarcity | RCP8.5 | | - | - | severe reducti on in water resourc es for about 8% of the global popula tion | severe reducti on in water resourc es for about 14% of the global popula tion | - | - | - | global |
| Haddeland et al., 2014 PNAS | Irrigatio n water scarcity | percentag e of populatio n under worsened water-stressed condition s based on Cumulati ve Abstracti on to Demand ratio | irrigation water scarcity increases with temperat ure in most regions | | | - | - | - | - | - | - | - | global |

1
2
3
4

Table SM7.2: literature considered in the expert judgement of risk transitions for figure 7.2

| Reference | Risk | variable (unit) | climate scenario | timeframe | GMS T level | Direction of impact | SSP 1 | SSP2 | SSP3 | SSP 4 | SSP5 | Region (Includin g Regional Differenc es) |
|------------------------|--|---|------------------|-----------|-------------|---|---|---|---------------------|-------|------|--|
| Food security | | | | | | | | | | | | |
| (Palazzo et al. 2017) | food availability | percent deviaton from 2010 Kilocalorie | RCP 8.5 | 2050 | | increase | up to 30% | | only up to 10% | | | West Africa |
| (Hasegawa et al. 2018) | change in crop yield combined with exposure and vulnerability based on prevalence of the undernourish ment (PoU) concept | populatio n at risk of hunger (million) | RCP2.6 | 2050 | | increasing population at risk of hunger | approx 2M | approx 5M | approx 24M | - | - | sub-Saharan Africa and South Asia have highest impacts |
| (Hasegawa et al. 2018) | change in crop yield combined with exposure and vulnerability based on prevalence of the undernourish ment (PoU) concept | populatio n at risk of hunger (million) | RCP6.0 | 2050 | | increasing population at risk of hunger | approx 5M (0-30M) (RCP to GMT conversio n based on SM SR15 ch3) | 24M (2-56M) (RCP to GMT conversio n based on SM SR15 ch3) | approx 80M (2-190M) | - | - | sub-Saharan Africa and South Asia have highest impacts |

| | | | | | | | | | | | | |
|-------------------------|-------------------|---|--|-----------------------|-----|---|---|--|---|---|---|--|
| (Byers et al. 2018) | crop yield change | crop yield change (Number of exposed and vulnerable people) | time sampling approach using a combination of RCPs | 2050 | 1.5 | | 2 | 8 | 20 | - | - | |
| (Byers et al. 2018) | crop yield change | crop yield change (Number of exposed and vulnerable people) | time sampling approach using a combination of RCPs | 2050 | 2 | | 24 | 81 | 178 | - | - | |
| (Byers et al. 2018) | crop yield change | crop yield change (Number of exposed and vulnerable people) | time sampling approach using a combination of RCPs | 2050 | 3 | | 118 | 406 | 854 | - | - | |
| (Wiebe et al. 2015) | Economic access | % change in price | RCP4.5 | 2050 | | Increase in price | ~3% to ~17% (interquartile range) | | | | | |
| (Wiebe et al. 2015) | Economic access | % change in price | RCP6.0 | 2050 | | Increase in price | - | 0 to ~12% increase (interquartile range) | - | - | - | |
| (Wiebe et al. 2015) | Economic access | % change in price | RCP8.5 | 2050 | | Increase in price | | ~5% to 30% (interquartile range), median by crop varies from 10% to 30%; restricting trade increases effects | | | | |
| (van Meijl et al. 2018) | Crop production | % change in production | RCP6.0 | 2050 | | Decrease in production | 2-3% decline | | 1-4% decline | | | |
| (van Meijl et al. 2018) | Economic access | % change in price | RCP6.0 | 2050 | | Increase in price | up to 5% | | up to 20% | | | |
| (Ishida et al. 2014) | undernourishment | DALYs attributable to childhood underweight (DAU) | Used RCP 4.5 for BAU | 2050 compared to 2005 | | generally decrease in undernourishment | Health burden decreases by 36.4 million DALYs by 2030 and to 11.6 DALYs by 2050 | decrease by 30.4 DALYs by 2030 and 17.0 DALYs by 2050 | decrease by 16.2 DALYs by 2030 but increase to 43.7 by 2050 | - | - | These are global statistics but there are regional differences. E.g. sub-Saharan Africa has higher DALYs |
| (Ishida et al. 2014) | undernourishment | DALYs attributable to childhood underweight (DAU) | Used RCP 2.6 | 2050 compared to 2005 | | generally decrease in undernourishment, although there are some climate impacts | Difference in health burden of 0.2% compared to BAU | Difference of 0.5% in 2050 compared to BAU | Difference of 2.0% compared to BAU | - | - | These are global statistics but there are regional differences. E.g. sub-Saharan Africa has higher DALYs |
| (Fujimori et al. 2018) | Economic access | GDP loss | RCP8.5 | 2100 | | Decline in GDP | 0% | 0.04% | 0.57% decrease | | | |

| | | | | | | | | | | | | |
|-------------------------|--|---|--|------|-----|---|--|--------------|--|---|---|---|
| | | | | | | | | | in "GDP change rate" | | | |
| (Springman et al. 2016) | Deaths due to changes in dietary and weight-related risk factors | Climate-related deaths | RCP2.6 to RCP8.5 | 2050 | | | more avoided deaths compared to SSP2 and 3 | intermediate | fewer avoided deaths | | | |
| Land degradation | | | | | | | | | | | | |
| (Byers et al. 2018) | habitat degradation | population (Million) exposed and vulnerable in relation to share of land area within a pixel being converted from natural land to agricultural land | time sampling approach using a combination of RCPs | 2050 | 1.5 | | 88 | 88 | 107 | - | - | non-drylands only; data provided by authors |
| (Byers et al. 2018) | habitat degradation | population (Million) exposed and vulnerable in relation to share of land area within a pixel being converted from natural land to agricultural land | time sampling approach using a combination of RCPs | 2050 | 2 | | 257 | 551 | 564 | - | - | non-drylands only; data provided by authors |
| (Byers et al. 2018) | habitat degradation | population (Million) exposed and vulnerable in relation to share of land area within a pixel being converted from natural land to agricultural land | time sampling approach using a combination of RCPs | 2050 | 3 | | 652 | 1068 | 1156 | - | - | non-drylands only; data provided by authors |
| (Hinkel et al. 2014) | flooding and sea level rise, Coastal erosion | number of people exposed to annual flooding | | 2100 | | | Lowest number of people flooded | - | highest number of people flooded | - | - | |
| (Hinkel et al. 2014) | Flood costs, Coastal erosion | cost of flooding (% GDP) | | 2100 | | The global costs of protecting the coast with dikes are | - | - | lowest costs under constant protection | - | highest costs under constant protection | |

| | | | | | | | | | | | |
|------------------------|-----------------------|---|--|------------------------|-----|--|---------|-----------|---|----|--|
| | | | | | | significant with annual investment and maintenance costs of US\$ 12–71 billion in 2100, but much smaller than the global cost of avoided damages even without accounting for indirect costs of damage to regional production supply. | | | but highest under enhanced protection ! | on | |
| (Zhang et al. 2018) | Extreme precipitation | population exposed to precipitation extremes (RX5day events exceeding 20-year return values) | time sampling approach on RCP8.5 and RCP4.5 | 2100 | 2 | exposed population steadily increases with temperature, with only marginal differences between SSPs | | | | | |
| (Knorr et al. 2016a) | fire | exposure (#people) | RCP4.5 transient | 2071-2100 vs 1971-2000 | 2 | | - | 560 | 646 | - | 508 globally |
| (Knorr et al. 2016a) | fire | exposure (#people) | RCP8.5 transient | 2071-2100 vs 1971-2000 | 4 | | - | 610 | 716 | - | 527 globally |
| (Knorr et al. 2016b) | fire | emissions (Pg C yr ⁻¹) | RCP4.5 transient | 2071-2100 vs 1971-2000 | 2 | | - | 1.22 | 1.11 | - | 1.31 globally |
| (Knorr et al. 2016b) | fire | emissions (Pg C yr ⁻¹) | RCP8.5 transient | 2071-2100 vs 1971-2000 | 4 | | - | 1.33 | 1.22 | - | 1.43 globally |
| Desertification | | | | | | | | | | | |
| (Zhang et al. 2018) | Extreme precipitation | population exposed to precipitation extremes (RX5day events exceeding 20-year return values) | time sampling approach on RCP8.5 and RCP4.5 | 2100 | 2 | exposed population steadily increases with temperature, with only marginal differences between SSPs | | | | | |
| (Byers et al. 2018) | water scarcity | water stress index (2050); population exposed and vulnerable in drylands (Units: Million and percentage of drylands population) | time sampling approach using a combination of RCPs | 2050 | 1.5 | | 76 (2%) | 349 (10%) | 783 (20%) | - | - Dryland only: data provided by authors |

| | | | | | | | | | | | | |
|--------------------------------|----------------|---|--|-----------|---|--|----------|------------|---|----------|----------|--|
| (Byers et al. 2018) | water scarcity | water stress index (2050); population exposed and vulnerable in drylands (Units: Million and percentage of drylands population) | time sampling approach using a combination of RCPs | 2050 | 2 | | 82 (3%) | 391 (11%) | 864 (22%) | - | - | Dryland only: data provided by authors |
| (Byers et al. 2018) | water scarcity | water stress index (2050); population exposed and vulnerable in drylands (Units: Million and percentage of drylands population) | time sampling approach using a combination of RCPs | 2050 | 3 | | 91 (3%) | 418 (12%) | 919 (24%) | - | - | Dryland only: data provided by authors |
| (Arnell and Lloyd-Hughes 2014) | water scarcity | Numbers of people (millions) exposed to increased water resources stress | RCP2.6 | 2050 | | | 379–2997 | 473–3434 | 626–4088 | 508–3481 | 418–3033 | |
| (Arnell and Lloyd-Hughes 2014) | water scarcity | Numbers of people (millions) exposed to increased water resources stress | RCP4.5 | 2050 | | | 810–2845 | 881–3239 | 1037–3975 | 884–3444 | 854–2879 | |
| (Arnell and Lloyd-Hughes 2014) | water scarcity | Numbers of people (millions) exposed to increased water resources stress | RCP6 | 2050 | | | 759–2668 | 807–3054 | 924–3564 | 809–3227 | 803–2682 | |
| (Arnell and Lloyd-Hughes 2014) | water scarcity | Numbers of people (millions) exposed to increased water resources stress | RCP8.5 | 2050 | | | 802–2947 | (919–3416) | 1006–4201 | 950–3519 | 854–2981 | |
| (Hanasaki et al. 2013) | water scarcity | Population living in grid cells with CAD < 0.5 | RCP8.5 | 2041–2070 | | | - | - | 4188 - 4434 (baseline is ~2000; all regions increase) | - | - | Global. Paper includes maps and graphs with regional |

| | | | | | | | | | | | | |
|------------------------|---|---|--------|-----------|--|--|---|--------------|---|---|---|--|
| (Hanasaki et al. 2013) | water scarcity | Population living in grid cells with CAD < 0.5 (millions) | RCP6.0 | 2041-2070 | | | 2853 - 3043 (baseline is ~2000; all regions increase) | | | | | information. Global Paper includes maps and graphs with regional information. |
| UNCCD, 2017 | mean species abundance, aridity; biodiversity, land degradation, water scarcity | population living in drylands | | | | | - | 43% increase | - | - | - | |

1
2**Table SM7.3: literature considered in the expert judgement of risk transitions for figure 7.3**

| Reference | Risk | Variable | Climate scenario | SSP | Timeframe | Non-climatic hazard | Bioenergy area | Impacts | Notes |
|-------------------------|------------------------------------|---|---|------|---|----------------------|----------------|---|-------|
| (Humpeöder et al. 2017) | trade-offs with SDGs | sustainability indicators: SDG 2; 7; 13; 14; 15 | no climate change (consistent with strong mitigation) | SSP1 | 2100 compared to baseline without bioenergy | bioenergy deployment | 636 Mha | only slight impact on sustainability indicators (i.e. no trade-offs due to lower food demand in SSP1) compared to baseline | |
| (Humpeöder et al. 2017) | trade-offs with SDGs | sustainability indicators: SDG 2; 7; 13; 14; 15 | no climate change (consistent with strong mitigation) | SSP2 | 2100 compared to baseline without bioenergy | bioenergy deployment | 636 Mha | pronounced decrease in all sustainability indicators (i.e. increase in adverse side-effects) compared to case without bioenergy | |
| (Humpeöder et al. 2017) | trade-offs with SDGs | sustainability indicators: SDG 2; 7; 13; 14; 15 | no climate change (consistent with strong mitigation) | SSP5 | 2100 compared to baseline without bioenergy | bioenergy deployment | 636 Mha | pronounced decrease in all sustainability indicators (i.e. increase in adverse side-effects) even more severe than in SSP2 | |
| (Heck et al. 2018) | planetary boundaries transgression | Planetary Boundaries (PBs): biosphere integrity; land-system change; biogeochemical flows; freshwater use | RCP2.6 | SSP1 | 2050 compared to baseline without bioenergy | bioenergy deployment | 870Mha | upper limit of most PBs is transgressed implying high risk of irreversible shifts | |
| (Heck et al. 2018) | planetary boundaries transgression | Planetary Boundaries (PBs): biosphere integrity; land-system change; biogeochemical flows; freshwater use | RCP2.6 | SSP2 | 2050 compared to baseline without bioenergy | bioenergy deployment | 778Mha | upper limit of most PBs is transgressed implying high risk of irreversible shifts | |
| (Boysen et al. 2017) | food production | kcal cap-1 day-1 production loss (%); N application (Mt yr ⁻¹) | 4.5°C trajectory | NA | 2100 | bioenergy deployment | 1078Mha | -43%; 96 Mt yr ⁻¹ | |
| (Boysen et al. 2017) | food production | kcal cap-1 day-1 production loss | 4.5°C trajectory | NA | 2100 | bioenergy deployment | 2176Mha | -73%; 151 Mt yr ⁻¹ | |

| | | | | | | | | | |
|----------------------------------|------------------------------|--|--|------|----------------------------|---|---|--------------------------------|--|
| | | (%); N application (Mt yr ⁻¹) | | | | | | | |
| (Boysen et al. 2017) | food production | kcal cap-1 day-1 production loss (%); N application (Mt yr ⁻¹) | 4.5°C trajectory | NA | 2100 | bioenergy deployment | 4267Mha | -100%; 196 Mt yr ⁻¹ | |
| (Hasegawa et al. 2018) | population at risk of hunger | population at risk of hunger (million) | RCP2.6 | SSP1 | 2050 compared to baseline | mitigation policies (including bioenergy) | 262Mha (106-490) (provided by authors) | approx +25M | |
| (Hasegawa et al. 2018) | population at risk of hunger | population at risk of hunger (million) | RCP2.6 | SSP2 | 2050 compared to baseline? | mitigation policies (including bioenergy) | 752Mha (175-1904) (provided by authors) | approx +78M (0-170) | |
| (Hasegawa et al. 2018) | population at risk of hunger | population at risk of hunger (million) | RCP2.6 | SSP3 | 2050 compared to baseline? | mitigation policies (including bioenergy) | 813Mha (171-1983) (provided by authors) | approx +120M | |
| (Fujimori et al. 2018) | population at risk of hunger | population at risk of hunger (million) | RCP2.6 | SSP1 | 2050 compared to baseline | mitigation policies (including bioenergy) | 90Mha | approx +20M | |
| (Fujimori et al. 2018) | population at risk of hunger | population at risk of hunger (million) | RCP2.6 | SSP2 | 2050 compared to baseline | mitigation policies (including bioenergy) | 170Mha | approx +100M | |
| (Fujimori et al. 2018) | population at risk of hunger | population at risk of hunger (million) | RCP2.6 | SSP3 | 2050 compared to baseline | mitigation policies (including bioenergy) | 220Mha | approx +260M | |
| (Obersteiner et al. 2016) | agricultural water use | km3 | | SSP1 | 2030 | bioenergy | 210Mha | approx + 13 km3 | |
| (Obersteiner et al. 2016) | agricultural water use | km3 | | SSP2 | 2030 | bioenergy | 210Mha | approx +12km3 | |
| (Obersteiner et al. 2016) | agricultural water use | km3 | | SSP3 | 2030 | bioenergy | 210Mha | approx +11km3 | |
| (Hejazi et al. 2014) | bioenergy water withdrawal | km3 | | SSP3 | 2050 | bioenergy | 150 Mha | approx +300km3 | Paper uses a precursor to the SSP3, with a similar population and storyline. |
| (Hasegawa et al. 2015) | population at risk of hunger | population | RCP2.6 | SSP2 | 2050 | bioenergy | 280Mha | approx +2M | |
| Fujimori et al., NSust, accepted | population at risk of hunger | population | No climate; but assessed in SM as small effect | SSP2 | 2050 | bioenergy | 38 - 395 Mha | approx 25 - 160 M | Difference between 1.5C scenario and Baseline for both bioenergy and impact. Total population at risk of hunger is ~300 to >500 million; total increase in population at risk of hunger is 50 to 320 M. Authors state that roughly half is attributed to bioenergy; those numbers are included here. |
| Fujimori et al., NSust, accepted | population at risk of hunger | population | No climate; but assessed in SM as small effect | SSP2 | 2050 | bioenergy | 43 - 225 Mha | approx 20 - 145 M | Difference between 2C scenario and Baseline for both bioenergy and impact. Total population at risk of hunger is ~290 to ~500 million; total increase in |

| | | | | | | | | | |
|--|--|--|--|--|--|--|--|--|---|
| | | | | | | | | | population at risk of hunger is 40 to 290 M. Authors state that roughly half is attributed to bioenergy; those numbers are included here. |
|--|--|--|--|--|--|--|--|--|---|

1
2
3
4
5
6
7
8

Table SM7.4. Risks thresholds for different components of desertification, land degradation and food security as a function of global mean surface temperature change relative to pre-industrial times. The confidence levels are defined according to the IPCC guidance note on consistent treatment of uncertainties (Mastrandrea et al., 2010). These data are used in Figure 7.1

| Component | Risk Transition | Global mean surface temperature change above pre-industrial levels °C | | Confidence |
|-------------------------|--------------------------|---|------|------------|
| | | Min | Max | |
| Low Latitude Crop Yield | Undetectable to Moderate | Min | 0.5 | high |
| | | Max | 0.7 | |
| | Moderate to High | Min | 1.2 | medium |
| | | Max | 2.2 | |
| | High to Very High | Min | 3.0 | medium |
| | | Max | 4.0 | |
| Food Supply Stability | Undetectable to Moderate | Min | 0.75 | high |
| | | Max | 0.85 | |
| | Moderate to High | Min | 0.9 | medium |
| | | Max | 1.4 | |
| | High to Very High | Min | 1.5 | medium |
| | | Max | 2.5 | |
| Permafrost Degradation | Undetectable to Moderate | Min | 0.3 | high |
| | | Max | 0.7 | |
| | Moderate to High | Min | 1.1 | high |
| | | Max | 1.5 | |
| | High to Very High | Min | 1.8 | medium |
| | | Max | 2.3 | |
| Vegetation Loss | Undetectable to Moderate | Min | 0.7 | high |
| | | Max | 1.0 | |
| | Moderate to High | Min | 1.6 | medium |
| | | Max | 2.6 | |
| | High to Very High | Min | 2.6 | medium |
| | | Max | 4.0 | |
| Coastal Degradation | Undetectable to Moderate | Min | 0.8 | high |
| | | Max | 1.05 | |
| | Moderate to High | Min | 1.1 | high |
| | | Max | 1.6 | |
| | High to Very High | Min | 1.8 | high |
| | | Max | | |

| | | | | | |
|-------------------|----------------------------|--------------------------|-----|--------|--------|
| | | Max | 2.7 | | |
| Soil Erosion | Undetectable to Moderate | Min | 0.8 | medium | |
| | | Max | 1.2 | | |
| | Moderate to High | Min | 2.0 | low | |
| | | Max | 3.5 | | |
| | High to Very High | Min | 4.0 | low | |
| | | Max | 6.0 | | |
| Fire | Undetectable to Moderate | Min | 0.7 | high | |
| | | Max | 1.0 | | |
| | Moderate to High | Min | 1.3 | medium | |
| | | Max | 1.7 | | |
| | High to Very High | Min | 2.5 | medium | |
| | | Max | 3.0 | | |
| | Water Scarcity in Drylands | Undetectable to Moderate | Min | 0.7 | high |
| | | | Max | 1.0 | |
| Moderate to High | | Min | 1.5 | medium | |
| | | Max | 2.5 | | |
| High to Very High | | Min | 2.5 | medium | |
| | | Max | 3.5 | | |
| Food Access | | Undetectable to Moderate | Min | 0.8 | medium |
| | | | Max | 1.1 | |
| | Moderate to High | Min | 1.4 | low | |
| | | Max | 2.4 | | |
| | High to Very High | Min | 2.4 | low | |
| | | Max | 3.4 | | |
| | Food Nutrition | Undetectable to Moderate | Min | 1.1 | low |
| | | | Max | 1.7 | |
| Moderate to High | | Min | 1.9 | low | |
| | | Max | 2.2 | | |
| High to Very High | | Min | 2.3 | low | |
| | | Max | 3.3 | | |

1

2

3

7. SM. 1. Additional embers

4

5 Details of two embers (nutrition and coastal degradation) where not included in Chapter 7 due to
6 space limitations. Changes in atmospheric CO₂, will result in reduced nutritional value of crops
7 (iron, protein, zinc, other micronutrients, and increases in mycotoxins), impacting food
8 utilization, with potential risks to health of vulnerable groups such as children and pregnant
9 women (*high confidence, high agreement*). This may create nutrition-related health risks for 600
10 million people (Zhou et al. 2018). Further details are provided in Chapter 5 of this Report.

11

12 Coastal flooding and degradation bring risk of damage to infrastructure and livelihoods. There are
13 very few global studies investigating past changes in coastal degradation (erosion and flooding)
14 and associated risk (Muis et al. 2018; Mentaschi et al. 2018), yet strong evidence exists that

1 anthropogenic climate change is already affecting the main drivers of coastal degradation,
 2 including: mean and extreme sea level (IPCC, 2013), storm surges (Wahl et al. 2015) and tropical
 3 cyclones (Kossin 2018). It is also clear that land-based processes, such as groundwater extraction
 4 and land subsidence, may impact coastal degradation {See Chapter 4, including 4.8.5}.

5
 6 At 1.5°C there is a high risk of destruction of coastal infrastructure and livelihoods (Hoegh-
 7 Guldberg et al. 2018) (*high confidence*). There is an associated strong increase in people and
 8 assets exposed to mean and extreme sea level rise and to coastal flooding above 1.5°C. Very high
 9 risks start to occur above 1.8 °C (*high confidence*) (Hanson et al. 2011; Vousdoukas et al. 2017;
 10 Jevrejeva et al. 2018; Hallegatte et al. 2013). Impacts of climate change on coasts is further
 11 explored in the IPCC Special Report on the Ocean and Cryosphere in a Changing Climate.

12 13 7. SM 2 SSP and Mitigation Burning Embers

14
 15 **Table SM7.5 Risks thresholds associated to desertification, land degradation and food security as a**
 16 **function of Global mean surface temperature change relative to pre-industrial levels and socio-**
 17 **economic development. Risks associated to desertification include, population exposed and**
 18 **vulnerable to water scarcity and changes in irrigation supply and demand. Risks related to land**
 19 **degradation include vegetation loss, population exposed to fire and floods, costs of floods, extent of**
 20 **deforestation, and ecosystem services including the ability of land to sequester carbon. Risks to food**
 21 **security include population at risk of hunger, food price increases, disability adjusted life years. The**
 22 **risks are assessed for two contrasted socio-economic futures (SSP1 and SSP3) under unmitigated**
 23 **climate change up to 3°C. These data are used in Figure 7.2.**
 24

| Component | Risk Transition | Global mean surface temperature change above pre-industrial levels °C | | Confidence |
|-------------------------|--------------------------|---|-----|-------------------------------|
| | | Min | Max | |
| Land Degradation (SSP1) | Undetectable to Moderate | Min | 0.7 | High |
| | | Max | 1.0 | |
| | Moderate to High | Min | 1.8 | low |
| | | Max | 2.8 | |
| | High to Very High | Min | | does not reach this threshold |
| | | Max | | |
| Land Degradation (SSP3) | Undetectable to Moderate | Min | 0.7 | High |
| | | Max | 1.0 | |
| | Moderate to High | Min | 1.4 | Medium |
| | | Max | 2.0 | |
| | High to Very High | Min | 2.2 | Medium |
| | | Max | 2.8 | |
| Food Security (SSP1) | Undetectable to Moderate | Min | 0.5 | Medium |
| | | Max | 1.0 | |
| | Moderate to High | Min | 2.5 | Medium |
| | | Max | 3.5 | |
| | High to Very High | Min | | does not reach this threshold |
| | | Max | | |

| | | | | |
|------------------------|--------------------------|-----|-----|-------------------------------|
| Food Security (SSP3) | Undetectable to Moderate | Min | 0.5 | Medium |
| | | Max | 1.0 | |
| | Moderate to High | Min | 1.3 | Medium |
| | | Max | 1.7 | |
| | High to Very High | Min | 2 | Medium |
| | | Max | 2.7 | |
| Desertification (SSP1) | Undetectable to Moderate | Min | 0.7 | High |
| | | Max | 1.0 | |
| | Moderate to High | Min | | Does not reach this threshold |
| | | Max | | |
| | High to Very High | Min | | Does not reach this threshold |
| | | Max | | |
| Desertification (SSP3) | Undetectable to Moderate | Min | 0.7 | High |
| | | Max | 1.0 | |
| | Moderate to High | Min | 1.2 | Medium |
| | | Max | 1.5 | |
| | High to Very High | Min | 1.5 | Medium |
| | | Max | 2.8 | |

1
2
3
4
5
6
7
8
9

Table SM7.6 Risk thresholds associated with 2nd generation bioenergy crop deployment (in 2050) as a land-based mitigation strategy under two SSPs (SSP1 and SSP3). The assessment is based on literature investigating the consequences of bioenergy expansion for food security, ecosystem loss and water scarcity, these indicators being aggregated as a single risk metric. These data are used in Figure 7.3.

| Component | Risk Transition | Land area used for bioenergy crop (Mkm ²) | | Confidence |
|---|--------------------------|---|-----|------------|
| Risk due to bioenergy deployment (SSP1) | Undetectable to Moderate | Min | 1 | Medium |
| | | Max | 4 | |
| | Moderate to High | Min | 6 | Low |
| | | Max | 8.7 | |
| | High to Very High | Min | 8.8 | Medium |
| | | Max | 20 | |
| Risk due to bioenergy deployment (SSP3) | Undetectable to Moderate | Min | 0.5 | Medium |
| | | Max | 1.5 | |
| | Moderate to High | Min | 1.5 | Low |
| | | Max | 3 | |
| | High to Very High | Min | 4 | Medium |
| | | Max | 8 | |

10

References

- 1
2
3 Abatzoglou, J. T., and A. P. Williams, 2016: Impact of anthropogenic climate change on wildfire
4 across western US forests. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1607171113.
5 ———, ———, and R. Barbero, 2019: Global Emergence of Anthropogenic Climate Change in Fire
6 Weather Indices. *Geophys. Res. Lett.*, **46**, 326–336, doi:10.1029/2018GL080959.
7 <http://doi.wiley.com/10.1029/2018GL080959> (Accessed April 14, 2019).
8 Aberman, N. L., and C. Tirado, 2014: Impacts of climate change on food utilization. *Global*
9 *Environmental Change*.
10 Abid, M., U. A. Schneider, and J. Scheffran, 2016: Adaptation to climate change and its impacts
11 on food productivity and crop income: Perspectives of farmers in rural Pakistan. *J. Rural*
12 *Stud.*, doi:10.1016/j.jrurstud.2016.08.005.
13 Andela, N., and Coauthors, 2017: A human-driven decline in global burned area. *Science*, **356**,
14 1356–1362, doi:10.1126/science.aal4108. <http://www.ncbi.nlm.nih.gov/pubmed/28663495>
15 (Accessed April 14, 2019).
16 Arnell, N. W., and B. Lloyd-Hughes, 2014: The global-scale impacts of climate change on water
17 resources and flooding under new climate and socio-economic scenarios. *Clim. Change*, **122**,
18 127–140, doi:10.1007/s10584-013-0948-4. [http://link.springer.com/10.1007/s10584-013-](http://link.springer.com/10.1007/s10584-013-0948-4)
19 [0948-4](http://link.springer.com/10.1007/s10584-013-0948-4) (Accessed April 11, 2019).
20 Asseng, S., and Coauthors, 2015: Rising temperatures reduce global wheat production. *Nat. Clim.*
21 *Chang.*, doi:10.1038/nclimate2470.
22 Asseng, S., D. Cammarano, B. Basso, U. Chung, P. D. Alderman, K. Sonder, M. Reynolds, and D.
23 B. Lobell, 2017: Hot spots of wheat yield decline with rising temperatures. *Glob. Chang.*
24 *Biol.*, doi:10.1111/gcb.13530.
25 Bellemare, M. F., 2015: Rising food prices, food price volatility, and social unrest. *Am. J. Agric.*
26 *Econ.*, doi:10.1093/ajae/aau038.
27 Bentz, B. J., and Coauthors, 2010: Climate Change and Bark Beetles of the Western United States
28 and Canada: Direct and Indirect Effects. *Bioscience*, **60**, 602–613,
29 doi:10.1525/bio.2010.60.8.6. [https://academic.oup.com/bioscience/article-](https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2010.60.8.6)
30 [lookup/doi/10.1525/bio.2010.60.8.6](https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2010.60.8.6) (Accessed April 14, 2019).
31 Betts, R. A., and Coauthors, 2018: Subject Areas : Author for correspondence : Changes in
32 climate extremes , fresh water availability and vulnerability to food insecurity projected at 1 .
33 5 ° C and 2 ° C global warming with a higher-resolution global climate model.
34 Black, R., W. N. Adger, N. W. Arnell, S. Dercon, A. Geddes, and D. Thomas, 2011: Migration
35 and global environmental change. *Global Environmental Change*.
36 Bojke, L., K. Claxton, Y. Bravo-Vergel, M. Sculpher, S. Palmer, and K. Abrams, 2010: Eliciting
37 distributions to populate decision analytic models. *Value Heal.*, **13**, 557–564.
38 Bonan, G. B., 2008: Forests and climate change: Forcings, feedbacks, and the climate benefits of
39 forests. *Science (80-.)*, **320**, 1444–1449, doi:10.1126/science.1155121.
40 Boysen, L. R., W. Lucht, D. Gerten, V. Heck, T. M. Lenton, and H. J. Schellnhuber, 2017: The
41 limits to global-warming mitigation by terrestrial carbon removal. *Earth's Futur.*, **5**, 463–
42 474, doi:10.1002/2016EF000469.
43 Burke, E. J., S. E. Chadburn, C. Huntingford, and C. D. Jones, 2018: CO₂ loss by permafrost
44 thawing implies additional emissions reductions to limit warming to 1.5 or 2 °C. *Environ.*
45 *Res. Lett.*, **13**, 024024, doi:10.1088/1748-9326/aaa138. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=2/a=024024?key=crossref.cf51ef78a9238c9ebf841752a904e4f6)
46 [9326/13/i=2/a=024024?key=crossref.cf51ef78a9238c9ebf841752a904e4f6](http://stacks.iop.org/1748-9326/13/i=2/a=024024?key=crossref.cf51ef78a9238c9ebf841752a904e4f6) (Accessed April
47 15, 2019).
48 Byers, E., and Coauthors, 2018: Global exposure and vulnerability to multi-sector development
49 and climate change hotspots. *Environ. Res. Lett.*, **13**, 055012, doi:10.1088/1748-
50 9326/aabf45. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=5/a=055012?key=crossref.dcb006b2e0b98d78e8d8ed4aa6eb51fb)
51 [9326/13/i=5/a=055012?key=crossref.dcb006b2e0b98d78e8d8ed4aa6eb51fb](http://stacks.iop.org/1748-9326/13/i=5/a=055012?key=crossref.dcb006b2e0b98d78e8d8ed4aa6eb51fb) (Accessed

- 1 April 11, 2019).
- 2 Calvin, K., M. Wise, P. Kyle, P. Patel, L. Clarke, and J. Edmonds, 2014: Trade-offs of different
3 land and bioenergy policies on the path to achieving climate targets. *Clim. Change*, **123**,
4 691–704, doi:10.1007/s10584-013-0897-y.
- 5 Chadburn, S. E., E. J. Burke, P. M. Cox, P. Friedlingstein, G. Hugelius, and S. Westermann,
6 2017a: An observation-based constraint on permafrost loss as a function of global warming.
7 *Nat. Clim. Chang.*, doi:10.1038/nclimate3262.
- 8 —, —, —, —, —, and —, 2017b: An observation-based constraint on permafrost
9 loss as a function of global warming. *Nat. Clim. Chang.*, **7**, 340–344,
10 doi:10.1038/nclimate3262. <http://www.nature.com/articles/nclimate3262> (Accessed April
11 15, 2019).
- 12 Challinor, A. J., J. Watson, D. B. Lobell, S. M. Howden, D. R. Smith, and N. Chhetri, 2014: A
13 meta-analysis of crop yield under climate change and adaptation. *Nat. Clim. Chang.*,
14 doi:10.1038/nclimate2153.
- 15 Challinor, A. J., W. N. Adger, and T. G. Benton, 2017: Climate risks across borders and scales.
16 *Nat. Clim. Chang.*, doi:10.1038/nclimate3380.
- 17 Chatzopoulos, T., I. Pérez Domínguez, M. Zampieri, and A. Toreti, 2019: Climate extremes and
18 agricultural commodity markets: A global economic analysis of regionally simulated events.
19 *Weather Clim. Extrem.*, doi:10.1016/j.wace.2019.100193.
- 20 Diffenbaugh, N. S., T. W. Hertel, M. Scherer, and M. Verma, 2012: Response of corn markets to
21 climate volatility under alternative energy futures. *Nat. Clim. Chang.*,
22 doi:10.1038/nclimate1491.
- 23 Donati, M., M. Zuppiroli, M. Riani, and G. Verga, 2016: The impact of investors in agricultural
24 commodity derivative markets. *Outlook Agric.*, doi:10.5367/oa.2016.0233.
- 25 Faye, B., and Coauthors, 2018: Impacts of 1.5 versus 2.0 °c on cereal yields in the West African
26 Sudan Savanna. *Environ. Res. Lett.*, doi:10.1088/1748-9326/aaab40.
- 27 Van der Fels-Klerx, H. J., C. Liu, and P. Battilani, 2016: Modelling climate change impacts on
28 mycotoxin contamination. *World Mycotoxin J.*, doi:10.3920/wmj2016.2066.
- 29 Fernandes, K., L. Verchot, W. Baethgen, V. Gutierrez-Velez, M. Pinedo-Vasquez, and C. Martius,
30 2017: Heightened fire probability in Indonesia in non-drought conditions: the effect of
31 increasing temperatures. *Environ. Res. Lett.*, **12**, 054002, doi:10.1088/1748-9326/aa6884.
32 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=5/a=054002?key=crossref.cb943b6a65dbd8384efb770f0ca7d8ed)
33 [9326/12/i=5/a=054002?key=crossref.cb943b6a65dbd8384efb770f0ca7d8ed](http://stacks.iop.org/1748-9326/12/i=5/a=054002?key=crossref.cb943b6a65dbd8384efb770f0ca7d8ed) (Accessed
34 April 14, 2019).
- 35 Fujimori, S., T. Hasegawa, J. Rogelj, X. Su, P. Havlik, V. Krey, K. Takahashi, and K. Riahi,
36 2018: Inclusive climate change mitigation and food security policy under 1.5 °C climate
37 goal. *Environ. Res. Lett.*, **13**, 074033, doi:10.1088/1748-9326/aad0f7.
38 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997)
39 [9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997](http://stacks.iop.org/1748-9326/13/i=7/a=074033?key=crossref.a32ea879418f0f13056428c7ab426997) (Accessed
40 April 10, 2019).
- 41 Gilbert, C. L., 2010: How to understand high food prices. *J. Agric. Econ.*, doi:10.1111/j.1477-
42 9552.2010.00248.x.
- 43 Gilbert, C. L., and C. W. Morgan, 2010: Food price volatility. *Philos. Trans. R. Soc. B Biol. Sci.*,
44 doi:10.1098/rstb.2010.0139.
- 45 Gosling, J., 2018: SHELFL: the Sheffield elicitation framework. *Elicitation*, Springer, Cham, 61–
46 93.
- 47 von der Gracht, H. A., 2012: Consensus measurement in Delphi studies. *Technol. Forecast. Soc.*
48 *Change*, **79**, 1524–1536.
- 49 Grant, S., M. Booth, and D. Khodyakov, 2018: Lack of preregistered analysis plans allows
50 unacceptable data mining for and selective reporting of consensus in Delphi studies. *J. Clin.*
51 *Epidemiol.*, **99**.

- 1 Haddeland, I., and Coauthors, 2014: Global water resources affected by human interventions and
2 climate change. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 3251–3256,
3 doi:10.1073/pnas.1222475110. <http://www.ncbi.nlm.nih.gov/pubmed/24344275> (Accessed
4 April 11, 2019).
- 5 Hallegatte, S., C. Green, R. J. Nicholls, and J. Corfee-Morlot, 2013: Future flood losses in major
6 coastal cities. *Nat. Clim. Chang.*, doi:10.1038/nclimate1979.
- 7 Hanasaki, N., and Coauthors, 2013: A global water scarcity assessment under Shared Socio-
8 economic Pathways – Part 2: Water availability and scarcity. *Hydrol.
9 Earth Syst. Sci.*, **17**, 2393–2413, doi:10.5194/hess-17-2393-2013. [https://www.hydrol-earth-
10 syst-sci.net/17/2393/2013/](https://www.hydrol-earth-syst-sci.net/17/2393/2013/) (Accessed April 11, 2019).
- 11 Hanson, S., R. Nicholls, N. Ranger, S. Hallegatte, J. Corfee-Morlot, C. Herweijer, and J. Chateau,
12 2011: A global ranking of port cities with high exposure to climate extremes. *Clim. Change*,
13 doi:10.1007/s10584-010-9977-4.
- 14 Harvey, C. A., Z. L. Rakotobe, N. S. Rao, R. Dave, H. Razafimahatratra, R. H. Rabarijohn, H.
15 Rajaofara, and J. L. MacKinnon, 2014: Extreme vulnerability of smallholder farmers to
16 agricultural risks and climate change in Madagascar. *Philos. Trans. R. Soc. B Biol. Sci.*,
17 doi:10.1098/rstb.2013.0089.
- 18 Hasegawa, T., S. Fujimori, K. Takahashi, and T. Masui, 2015: Scenarios for the risk of hunger in
19 the twenty-first century using Shared Socioeconomic Pathways. *Environ. Res. Lett.*, **10**,
20 14010.
- 21 ———, and Coauthors, 2018: Risk of increased food insecurity under stringent global climate
22 change mitigation policy. *Nat. Clim. Chang.*, **8**, 699–703, doi:10.1038/s41558-018-0230-x.
23 <http://www.nature.com/articles/s41558-018-0230-x> (Accessed April 10, 2019).
- 24 Hasson, F., and S. Keeney, 2011: Enhancing rigour in the Delphi technique research. *Technol.
25 Forecast. Soc. Change*, **78**, 1695–1704.
- 26 Headey, D., 2011: Rethinking the global food crisis: The role of trade shocks. *Food Policy*,
27 doi:10.1016/j.foodpol.2010.10.003.
- 28 Heck, V., D. Gerten, W. Lucht, and A. Popp, 2018: Biomass-based negative emissions difficult to
29 reconcile with planetary boundaries. *Nat. Clim. Chang.*, **8**, 151–155, doi:10.1038/s41558-
30 017-0064-y. <http://www.nature.com/articles/s41558-017-0064-y> (Accessed April 10, 2019).
- 31 Hejazi, M. I., and Coauthors, 2014: Integrated assessment of global water scarcity over the 21st
32 century under multiple climate change mitigation policies. *Hydrol. Earth Syst. Sci.*, **18**,
33 2859–2883, doi:10.5194/hess-18-2859-2014. [https://www.hydrol-earth-syst-
34 sci.net/18/2859/2014/](https://www.hydrol-earth-syst-sci.net/18/2859/2014/) (Accessed April 10, 2019).
- 35 Hember, R. A., W. A. Kurz, and N. C. Coops, 2017: Relationships between individual-tree
36 mortality and water-balance variables indicate positive trends in water stress-induced tree
37 mortality across North America. *Glob. Chang. Biol.*, **23**, 1691–1710, doi:10.1111/gcb.13428.
38 <http://doi.wiley.com/10.1111/gcb.13428> (Accessed April 14, 2019).
- 39 Hertel, T. W., M. B. Burke, and D. B. Lobell, 2010: The poverty implications of climate-induced
40 crop yield changes by 2030. *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2010.07.001.
- 41 Hinkel, J., and Coauthors, 2014: Coastal flood damage and adaptation costs under 21st century
42 sea-level rise. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1222469111.
- 43 Hjort, J., O. Karjalainen, J. Aalto, S. Westermann, V. E. Romanovsky, F. E. Nelson, B.
44 Etzelmüller, and M. Luoto, 2018: Degrading permafrost puts Arctic infrastructure at risk by
45 mid-century. *Nat. Commun.*, **9**, 5147, doi:10.1038/s41467-018-07557-4.
- 46 Hoegh-Guldberg, O., and Coauthors, 2018: Impacts of 1.5°C global warming on natural and
47 human systems. *Global Warming of 1.5 °C an IPCC special report on the impacts of global
48 warming of 1.5 °C above pre-industrial levels and related global greenhouse gas emission
49 pathways, in the context of strengthening the global response to the threat of climate change*
50 http://report.ipcc.ch/sr15/pdf/sr15_chapter3.pdf (Accessed October 29, 2018).
- 51 Huang, J., H. Yu, A. Dai, Y. Wei, and L. Kang, 2017: Drylands face potential threat under 2 °C

- 1 global warming target. *Nat. Clim. Chang.*, **7**, 417–422, doi:10.1038/nclimate3275.
2 <http://www.nature.com/articles/nclimate3275> (Accessed April 11, 2019).
- 3 Humpenöder, F., and Coauthors, 2017: Large-scale bioenergy production: How to resolve
4 sustainability trade-offs? *Environ. Res. Lett.*, doi:10.1088/1748-9326/aa9e3b.
5 <http://iopscience.iop.org/article/10.1088/1748-9326/aa9e3b> (Accessed December 8, 2017).
- 6 IPCC, 2018: Summary for Policy Makers. *IPCC Special Report on the impacts of global*
7 *warming of 1.5°C*, William Solecki http://report.ipcc.ch/sr15/pdf/sr15_spm_final.pdf
8 (Accessed November 1, 2018).
- 9 IPCC 2014, 2014: *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II*
10 *and III to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change.*
11 IPCC, Ed. Gian-Kasper Plattner, Geneva, <http://www.ipcc.ch>. (Accessed October 31, 2018).
- 12 Ishida, H., and Coauthors, 2014: Global-scale projection and its sensitivity analysis of the health
13 burden attributable to childhood undernutrition under the latest scenario framework for
14 climate change research. *Environ. Res. Lett.*, **9**, 064014, doi:10.1088/1748-9326/9/6/064014.
15 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/9/i=6/a=064014?key=crossref.288e05e1c145e8fa5c1873d6d19c9811)
16 [9326/9/i=6/a=064014?key=crossref.288e05e1c145e8fa5c1873d6d19c9811](http://stacks.iop.org/1748-9326/9/i=6/a=064014?key=crossref.288e05e1c145e8fa5c1873d6d19c9811) (Accessed April
17 14, 2019).
- 18 Jevrejeva, S., L. P. Jackson, A. Grinsted, D. Lincke, and B. Marzeion, 2018: Flood damage costs
19 under the sea level rise with warming of 1.5°C and 2°C. *Environ. Res. Lett.*,
20 doi:10.1088/1748-9326/aacc76.
- 21 Jolly, W. M., M. A. Cochrane, P. H. Freeborn, Z. A. Holden, T. J. Brown, G. J. Williamson, and
22 D. M. J. S. Bowman, 2015: Climate-induced variations in global wildfire danger from 1979
23 to 2013. *Nat. Commun.*, **6**, 7537, doi:10.1038/ncomms8537.
24 <http://www.nature.com/articles/ncomms8537> (Accessed April 14, 2019).
- 25 Kelley, C., S. Mohtadi, M. Cane, R. Seager, and Y. Kushnir, 2017: Commentary on the Syria
26 case: Climate as a contributing factor. *Polit. Geogr.*, doi:10.1016/j.polgeo.2017.06.013.
- 27 Knorr, W., A. Arneth, and L. Jiang, 2016a: Demographic controls of future global fire risk. *Nat.*
28 *Clim. Chang.*, **6**, 781–785, doi:10.1038/nclimate2999.
29 <http://www.nature.com/articles/nclimate2999> (Accessed April 12, 2019).
- 30 —, L. Jiang, and A. Arneth, 2016b: Climate, CO₂ and human population impacts on global wildfire emissions. *Biogeosciences*, **13**, 267–282,
31 doi:10.5194/bg-13-267-2016. <https://www.biogeosciences.net/13/267/2016/> (Accessed
32 April 21, 2019).
- 34 Kossin, J. P., 2018: A global slowdown of tropical-cyclone translation speed. *Nature*, **558**, 104–
35 107, doi:10.1038/s41586-018-0158-3. <http://dx.doi.org/10.1038/s41586-018-0158-3>.
- 36 Kreidenweis, U., F. Humpenöder, M. Stevanović, B. L. Bodirsky, E. Kriegler, H. Lotze-Campen,
37 and A. Popp, 2016: Afforestation to mitigate climate change: impacts on food prices under
38 consideration of albedo effects. *Environ. Res. Lett.*, **11**, 085001, doi:10.1088/1748-
39 9326/11/8/085001. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=8/a=085001?key=crossref.498dab12c59b27f71805e8cdbafc36f1)
40 [9326/11/i=8/a=085001?key=crossref.498dab12c59b27f71805e8cdbafc36f1](http://stacks.iop.org/1748-9326/11/i=8/a=085001?key=crossref.498dab12c59b27f71805e8cdbafc36f1) (Accessed
41 December 9, 2017).
- 42 Li, Z., and H. Fang, 2016: Impacts of climate change on water erosion: A review. *Earth-Science*
43 *Rev.*, **163**, 94–117, doi:10.1016/J.EARSCIREV.2016.10.004.
44 <https://www.sciencedirect.com/science/article/pii/S0012825216303555> (Accessed April 14,
45 2019).
- 46 Magan, N., A. Medina, and D. Aldred, 2011: Possible climate-change effects on mycotoxin
47 contamination of food crops pre- and postharvest. *Plant Pathol.*, doi:10.1111/j.1365-
48 3059.2010.02412.x.
- 49 Marchand, P., and Coauthors, 2016: Reserves and trade jointly determine exposure to food supply
50 shocks. *Environ. Res. Lett.*, doi:10.1088/1748-9326/11/9/095009.
- 51 Medina-Elizalde, M., and E. J. Rohling, 2012: Collapse of classic maya civilization related to

- 1 modest reduction in precipitation. *Science* (80-.), doi:10.1126/science.1216629.
- 2 Medina, A., A. Akbar, A. Baazeem, A. Rodriguez, and N. Magan, 2017: Climate change, food
3 security and mycotoxins: Do we know enough? *Fungal Biol. Rev.*,
4 doi:10.1016/j.fbr.2017.04.002.
- 5 van Meijl, H., and Coauthors, 2018: Comparing impacts of climate change and mitigation on
6 global agriculture by 2050. *Environ. Res. Lett.*, **13**, 064021, doi:10.1088/1748-9326/aabdc4.
7 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=6/a=064021?key=crossref.42a4eb1897f2ed545f2b0dc439d03e64)
8 [9326/13/i=6/a=064021?key=crossref.42a4eb1897f2ed545f2b0dc439d03e64](http://stacks.iop.org/1748-9326/13/i=6/a=064021?key=crossref.42a4eb1897f2ed545f2b0dc439d03e64) (Accessed
9 April 12, 2019).
- 10 Mentaschi, L., M. I. Voudoukas, J. F. Pekel, E. Voukouvalas, and L. Feyen, 2018: Global long-
11 term observations of coastal erosion and accretion. *Sci. Rep.*, doi:10.1038/s41598-018-
12 30904-w.
- 13 Moretti, A., M. Pascale, and A. F. Logrieco, 2018: Mycotoxin risks under a climate change
14 scenario in Europe. *Trends in Food Science and Technology*.
- 15 Morris, G. P., S. Reis, S. A. Beck, L. E. Fleming, W. N. Adger, T. G. Benton, and M. H.
16 Depledge, 2017: Scoping the proximal and distal dimensions of climate change on health
17 and wellbeing. *Environ. Heal. A Glob. Access Sci. Source*, doi:10.1186/s12940-017-0329-y.
- 18 Muis, S., I. D. Haigh, and J. C. J. H. Aerts, 2018: Earth ' s Future In fl uence of El Niño-Southern
19 Oscillation on Global Coastal Flooding Earth ' s Future. 1311–1322,
20 doi:10.1029/2018EF000909.
- 21 Mukherjee, N., J. Hugé, W. J. Sutherland, J. Mcneill, M. Van Opstal, F. Dahdouh-Guebas, and N.
22 Koedam, 2015: The Delphi technique in ecology and biological conservation: Applications
23 and guidelines. *Methods Ecol. Evol.*, **6**, 1097–1109, doi:10.1111/2041-210X.12387.
- 24 Myers, S. S., and Coauthors, 2014: Increasing CO2 threatens human nutrition. *Nature*,
25 doi:10.1038/nature13179.
- 26 —, K. R. Wessells, I. Kloog, A. Zanobetti, and J. Schwartz, 2015: Effect of increased
27 concentrations of atmospheric carbon dioxide on the global threat of zinc deficiency: A
28 modelling study. *Lancet Glob. Heal.*, doi:10.1016/S2214-109X(15)00093-5.
- 29 Neumann, B., A. T. Vafeidis, J. Zimmermann, and R. J. Nicholls, 2015: Future coastal population
30 growth and exposure to sea-level rise and coastal flooding-a global assessment. *PLoS One*,
31 **10**, e0118571.
- 32 Nicholls, R. J., and A. Cazenave, 2010: Sea-level rise and its impact on coastal zones. *Science*
33 (80-.), **328**, 1517–1520.
- 34 O'Neill, B. C., and Coauthors, 2017: IPCC reasons for concern regarding climate change risks.
35 *Nat. Clim. Chang.*, **7**, 28–37, doi:10.1038/nclimate3179.
- 36 Oakley, J. E., and A. O'Hagan, 2016: SHELF: the Sheffield Elicitation Framework (version 3.0).
37 <http://tonyohagan.co.uk/shelf>.
- 38 Obersteiner, M., and Coauthors, 2016: Assessing the land resource–food price nexus of the
39 Sustainable Development Goals. *Sci. Adv.*, **2**, e1501499, doi:10.1126/sciadv.1501499.
40 <http://advances.sciencemag.org/lookup/doi/10.1126/sciadv.1501499> (Accessed April 21,
41 2019).
- 42 Palazzo, A., and Coauthors, 2017: Linking regional stakeholder scenarios and shared
43 socioeconomic pathways: Quantified West African food and climate futures in a global
44 context. *Glob. Environ. Chang.*, **45**, 227–242, doi:10.1016/J.GLOENVCHA.2016.12.002.
45 <https://www.sciencedirect.com/science/article/pii/S0959378016305751> (Accessed April 21,
46 2019).
- 47 Parry, M. L., C. Rosenzweig, A. Iglesias, M. Livermore, and G. Fischer, 2004: Effects of climate
48 change on global food production under SRES emissions and socio-economic scenarios.
49 *Glob. Environ. Chang.*, doi:10.1016/j.gloenvcha.2003.10.008.
- 50 Paterson, R. R. M., and N. Lima, 2011: Further mycotoxin effects from climate change. *Food Res.*
51 *Int.*, doi:10.1016/j.foodres.2011.05.038.

- 1 Porter, J. R., L. Xie, A. J. Challinor, K. Cochrane, S. M. Howden, M. M. Iqbal, D. B. Lobell, and
2 M. I. Travasso, 2014: Food security and food production systems. *Climate Change 2014:*
3 *Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution*
4 *of Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on*
5 *Climate Change*, 485–533.
- 6 ———, and Coauthors, 2015: Food security and food production systems. *Climate Change 2014*
7 *Impacts, Adaptation and Vulnerability: Part A: Global and Sectoral Aspects*.
- 8 Rivera-Ferre, M. G., and Coauthors, 2016: Local agriculture traditional knowledge to ensure food
9 availability in a changing climate: revisiting water management practices in the Indo-
10 Gangetic Plains. *Agroecol. Sustain. Food Syst.*, doi:10.1080/21683565.2016.1215368.
- 11 Roberts, M. J., and W. Schlenker, 2013: Identifying supply and demand elasticities of agricultural
12 commodities: Implications for the US ethanol mandate. *Am. Econ. Rev.*,
13 doi:10.1257/aer.103.6.2265.
- 14 Rosenzweig, C., and Coauthors, 2014: Assessing agricultural risks of climate change in the 21st
15 century in a global gridded crop model intercomparison. *Proc. Natl. Acad. Sci.*,
16 doi:10.1073/pnas.1222463110.
- 17 Salmon, J. M., M. A. Friedl, S. Froking, D. Wisser, and E. M. Douglas, 2015: Global rain-fed,
18 irrigated, and paddy croplands: A new high resolution map derived from remote sensing,
19 crop inventories and climate data. *Int. J. Appl. Earth Obs. Geoinf.*,
20 doi:10.1016/j.jag.2015.01.014.
- 21 Schleussner, C. F., and Coauthors, 2016: Differential climate impacts for policy-relevant limits to
22 global warming: The case of 1.5 °c and 2 °c. *Earth Syst. Dyn.*, doi:10.5194/esd-7-327-2016.
- 23 Schmidhuber, J., and F. N. Tubiello, 2007: Global food security under climate change. *Proc. Natl.*
24 *Acad. Sci.*, doi:10.1073/pnas.0701976104.
- 25 Smith, M. R., C. D. Golden, and S. S. Myers, 2017: Potential rise in iron deficiency due to future
26 anthropogenic carbon dioxide emissions. *GeoHealth*, doi:10.1002/2016gh000018.
- 27 Smith, P., R. S. Haszeldine, and S. M. Smith, 2016: Preliminary assessment of the potential for,
28 and limitations to, terrestrial negative emission technologies in the UK. *Environ. Sci.*
29 *Process. Impacts*, doi:10.1039/c6em00386a.
- 30 Sperber, D., D. Mortimer, P. Lorgelly, and D. Berlowitz, 2013: An expert on every street corner?
31 Methods for eliciting distributions in geographically dispersed opinion pools. *Value Heal.*,
32 **16**, 434–437.
- 33 Springmann, M., H. C. J. Godfray, M. Rayner, and P. Scarborough, 2016: Analysis and valuation
34 of the health and climate change cobenefits of dietary change. *Proc. Natl. Acad. Sci.*,
35 doi:10.1073/pnas.1523119113.
- 36 Sternberg, T., 2012: Chinese drought, bread and the Arab Spring. *Appl. Geogr.*,
37 doi:10.1016/j.apgeog.2012.02.004.
- 38 Sturrock, R. N., S. J. Frankel, A. V. Brown, P. E. Hennon, J. T. Kliejunas, K. J. Lewis, J. J.
39 Worrall, and A. J. Woods, 2011: Climate change and forest diseases. *Plant Pathol.*, **60**,
40 133–149, doi:10.1111/j.1365-3059.2010.02406.x.
- 41 Thompson, B., and M. J. Cohen, 2012: *The impact of climate change and bioenergy on nutrition.*
- 42 Tigchelaar, M., D. Battisti, R. Naylor, and D. Ray, 2018: Probability of globally synchronized
43 maize production shocks. *Proc. Natl. Acad. Sci.*, **115**, 6644–6649.
- 44 Tirado, M. C., R. Clarke, L. A. Jaykus, A. McQuatters-Gollop, and J. M. Frank, 2010: Climate
45 change and food safety: A review. *Food Res. Int.*, doi:10.1016/j.foodres.2010.07.003.
- 46 Vanmaercke, M., and Coauthors, 2016: How fast do gully headcuts retreat? *Earth-Science Rev.*,
47 **154**, 336–355, doi:10.1016/J.EARSCIREV.2016.01.009.
48 <https://www.sciencedirect.com/science/article/pii/S0012825216300083> (Accessed April 14,
49 2019).
- 50 Verma, M., T. Hertel, and N. Diffenbaugh, 2014: Market-oriented ethanol and corn-trade policies
51 can reduce climate-induced US corn price volatility. *Environ. Res. Lett.*, doi:10.1088/1748-

- 1 9326/9/6/064028.
- 2 Vermeulen, S. J., and Coauthors, 2012: Options for support to agriculture and food security under
3 climate change. *Environ. Sci. Policy*, doi:10.1016/j.envsci.2011.09.003.
- 4 Vousdoukas, M. I., L. Mentaschi, E. Voukouvalas, M. Verlaan, and L. Feyen, 2017: Extreme sea
5 levels on the rise along Europe’s coasts. *Earth’s Futur.*, doi:10.1002/2016EF000505.
- 6 —, —, —, —, S. Jevrejeva, L. P. Jackson, and L. Feyen, 2018: Global probabilistic
7 projections of extreme sea levels show intensification of coastal flood hazard. *Nat.*
8 *Commun.*, **9**, 2360, doi:10.1038/s41467-018-04692-w.
9 <http://www.nature.com/articles/s41467-018-04692-w> (Accessed April 15, 2019).
- 10 Wahl, T., S. Jain, J. Bender, S. D. Meyers, and M. E. Luther, 2015: Increasing risk of compound
11 flooding from storm surge and rainfall for major US cities. *Nat. Clim. Chang.*, **5**, 1093–1097,
12 doi:10.1038/nclimate2736.
- 13 Westerling, A. L., H. G. Hidalgo, D. R. Cayan, and T. W. Swetnam, 2006: Warming and Earlier
14 Spring Increase Western U.S. Forest Wildfire Activity. *Science (80-.)*, **313**, 940–943,
15 doi:10.1126/SCIENCE.1128834.
- 16 Wiebe, K., and Coauthors, 2015: Climate change impacts on agriculture in 2050 under a range of
17 plausible socioeconomic and emissions scenarios. *Environ. Res. Lett.*, **10**, 085010,
18 doi:10.1088/1748-9326/10/8/085010. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b)
19 [9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b](http://stacks.iop.org/1748-9326/10/i=8/a=085010?key=crossref.acb559d1aa179071d5d2466fd63ceb3b) (Accessed
20 April 12, 2019).
- 21 Willenbockel, D., 2012: *Extreme weather events and crop price spikes in a changing climate.*
22 *Illustrative global simulation scenarios.*
- 23 Yang, J., H. Tian, B. Tao, W. Ren, J. Kush, Y. Liu, and Y. Wang, 2014: Spatial and temporal
24 patterns of global burned area in response to anthropogenic and environmental factors:
25 Reconstructing global fire history for the 20th and early 21st centuries. *J. Geophys. Res.*
26 *Biogeosciences*, **119**, 249–263, doi:10.1002/2013JG002532.
27 <http://doi.wiley.com/10.1002/2013JG002532> (Accessed April 14, 2019).
- 28 Zampieri, M., A. Ceglar, F. Dentener, and A. Toreti, 2017: Wheat yield loss attributable to heat
29 waves, drought and water excess at the global, national and subnational scales. *Environ. Res.*
30 *Lett.*, doi:10.1088/1748-9326/aa723b.
- 31 Zhang, W., T. Zhou, L. Zou, L. Zhang, and X. Chen, 2018: Reduced exposure to extreme
32 precipitation from 0.5 °C less warming in global land monsoon regions. *Nat. Commun.*, **9**,
33 3153, doi:10.1038/s41467-018-05633-3. [http://www.nature.com/articles/s41467-018-05633-](http://www.nature.com/articles/s41467-018-05633-3)
34 [3](http://www.nature.com/articles/s41467-018-05633-3) (Accessed April 12, 2019).
- 35 Zhao, C., and Coauthors, 2017: Temperature increase reduces global yields of major crops in four
36 independent estimates. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1701762114.
- 37 Zheng, J., L. Xiao, X. Fang, Z. Hao, Q. Ge, and B. Li, 2014: How climate change impacted the
38 collapse of the Ming dynasty. *Clim. Change*, doi:10.1007/s10584-014-1244-7.
- 39 Zimmermann, A., H. Webber, G. Zhao, F. Ewert, J. Kros, J. Wolf, W. Britz, and W. de Vries,
40 2017: Climate change impacts on crop yields, land use and environment in response to crop
41 sowing dates and thermal time requirements. *Agric. Syst.*, doi:10.1016/j.agsy.2017.07.007.
- 42 Zscheischler, J., and Coauthors, 2018: Future climate risk from compound events. *Nat. Clim.*
43 *Chang.*, doi:10.1038/s41558-018-0156-3.
- 44 2012: *The Russian Federation forest sector: outlook study to 2030.* FAO,
45 <http://agris.fao.org/agris-search/search.do?recordID=XF2013001279> (Accessed April 16,
46 2019).
- 47

Glossary

Coordinating Editor:

Renée van Diemen (Netherlands/UK)

Editorial Team:

Tim Benton (UK), Eduardo Calvo (Peru), Annette Cowie (Australia), Valérie Masson-Delmotte (France), Aziz Elbehri (Morocco), Karlheinz Erb (Italy), Giacomo Grassi (Italy), J.B. Robin Matthews (UK), Hans-Otto Pörtner (Germany), Andy Reisinger (New Zealand), Debra Roberts (South Africa), Priyadarshi Shukla (India), Andrey Sirin (Russian Federation), Jim Skea (UK), Murat Türkeş (Turkey), Nora M. Weyer (Germany), Sumaya Zakieldeen (Sudan), Panmao Zhai (China)

Notes:

Note that subterms are in italics beneath main terms

This glossary defines some specific terms as the Lead Authors intend them to be interpreted in the context of this report. Italicised words indicate that the term is defined in the Glossary.

1.5°C pathway

See *Pathways*.

2030 Agenda for Sustainable Development

A UN resolution in September 2015 adopting a plan of action for people, planet and prosperity in a new global development framework anchored in 17 *Sustainable Development Goals* (UN, 2015).

Acceptability of policy or system change

The extent to which a policy or system change is evaluated unfavourably or favourably, or rejected or supported, by members of the general public (public acceptability) or politicians or governments (political acceptability). Acceptability may vary from totally unacceptable/fully rejected to totally acceptable/fully supported; individuals may differ in how acceptable policies or system changes are believed to be.

Acclimatization

A change in functional or morphological traits occurring once or repeatedly (e.g., seasonally) during the lifetime of an individual organism in its natural environment. Through acclimatization the individual maintains performance across a range of environmental conditions. For a clear differentiation between findings in laboratory and field studies, the term acclimation is used in ecophysiology for the respective phenomena when observed in well-defined experimental settings. The term (adaptive) plasticity characterises the generally limited scope of changes in phenotype that an individual can reach through the process of acclimatization.

Activity

A practice or ensemble of practices that take place on a delineated area over a given period of time.

Activity data

Data on the magnitude of a human *activity* resulting in emissions or removals taking place during a given period of time. In the *Agriculture, Forestry and Other Land Use (AFOLU)* sector, data on area of different *land uses*, management systems, animal numbers, lime and fertiliser use are examples of activity data.

Adaptability

See *Adaptive capacity*.

Adaptation

In *human systems*, the process of adjustment to actual or expected *climate* and its effects, in order to moderate harm or exploit beneficial opportunities. In natural systems, the process of adjustment to actual climate and its effects; human intervention may facilitate adjustment to expected climate and its effects.

Incremental adaptation

Adaptation that maintains the essence and integrity of a system or process at a given scale (Park et al., 2012).

Transformational adaptation

Adaptation that changes the fundamental attributes of a *social-ecological system* in anticipation of *climate change* and its *impacts*.

Adaptation limits

The point at which an actor's objectives (or system needs) cannot be secured from intolerable risks through adaptive actions.

- Hard adaptation limit - No adaptive actions are possible to avoid intolerable risks.
- Soft adaptation limit - Options are currently not available to avoid intolerable risks through adaptive action.

See also *Adaptation options*, *Adaptive capacity*, and *Maladaptive actions (Maladaptation)*.

Adaptation behaviour

See *Human behaviour*.

Adaptation limits

See *Adaptation*.

Adaptation options

The array of strategies and measures that are available and appropriate for addressing *adaptation*. They include a wide range of actions that can be categorised as structural, institutional, ecological or behavioural.

See also *Adaptive capacity*, and *Maladaptive actions (Maladaptation)*.

Adaptation pathways

See *Pathways*.

Adaptive capacity

The ability of systems, *institutions*, humans and other organisms to adjust to potential damage, to take advantage of opportunities, or to respond to consequences (IPCC, 2014; MA, 2005).

See also *Adaptation*, *Adaptation options*, and *Maladaptive actions (Maladaptation)*.

Adaptive governance

See *Governance*.

Adverse side-effect

The negative effects that a policy or measure aimed at one objective might have on other objectives, without yet evaluating the net effect on overall social welfare. Adverse side-effects are often subject to uncertainty and depend on, among others, local circumstances and implementation practices.

See also *Co-benefits* and *Risk*.

Aerosol

A suspension of airborne solid or liquid particles, with a typical size between a few nanometres and 10 µm that reside in the *atmosphere* for at least several hours. The term aerosol, which includes both

the particles and the suspending gas, is often used in this report in its plural form to mean aerosol particles. Aerosols may be of either natural or *anthropogenic* origin. Aerosols may influence *climate* in several ways: through both interactions that scatter and/or absorb radiation and through interactions with cloud microphysics and other cloud properties, or upon deposition on snow or ice covered surfaces thereby altering their *albedo* and contributing to *climate feedback*. Atmospheric aerosols, whether natural or anthropogenic, originate from two different pathways: emissions of primary particulate matter (PM), and formation of secondary PM from gaseous *precursors*. The bulk of aerosols are of natural origin. Some scientists use group labels that refer to the chemical composition, namely: sea salt, organic carbon, *black carbon* (BC), mineral species (mainly desert dust), sulphate, nitrate, and ammonium. These labels are, however, imperfect as aerosols combine particles to create complex mixtures.

See also *Short-lived climate forcers (SLCF)*.

Afforestation

Conversion to *forest* of land that historically has not contained forests.

[Note: For a discussion of the term forest and related terms such as afforestation, *reforestation* and *deforestation*, in the context of reporting and accounting Article 3.3 and 3.4 activities under the *Kyoto Protocol*, see 2013 Revised Supplementary Methods and Good Practice Guidance Arising from the Kyoto Protocol.]

See also *Reforestation, Deforestation, Forest and Reducing Emissions from Deforestation and Forest Degradation (REDD+)*.

Agreement

In this report, the degree of agreement within the scientific body of knowledge on a particular finding is assessed based on multiple lines of *evidence* (e.g., mechanistic understanding, theory, data, models, expert judgement) and expressed qualitatively (Mastrandrea et al., 2010).

See also *Confidence, Likelihood, and Uncertainty*.

Agriculture, Forestry and Other Land Use (AFOLU)

In the context of national *greenhouse gas (GHG)* inventories under the *United Nations Convention on Climate Change (UNFCCC)*, AFOLU is the sum of the GHG inventory sectors Agriculture and Land Use, Land-Use Change and Forestry (LULUCF); see the 2006 IPCC Guidelines for National GHG Inventories for details. Given the difference in estimating the ‘anthropogenic’ *carbon dioxide (CO₂)* removals between countries and the global modelling community, the land-related net GHG emissions from global models included in this report are not necessarily directly comparable with LULUCF estimates in national GHG Inventories.

See also *Land-use change (LUC)* and *Land Use, Land-use change and Forestry (LULUCF)*.

Agrobiodiversity

‘The variety and variability of animals, plants and micro-organisms that are used directly or indirectly for food and agriculture, including crops, livestock, forestry and fisheries. It comprises the diversity of genetic resources (varieties, breeds) and species used for food, fodder, fibre, fuel and pharmaceuticals. It also includes the diversity of non-harvested species that support production (soil micro-organisms, predators, pollinators), and those in the wider environment that support agro-ecosystems (agricultural, pastoral, forest and aquatic) as well as the diversity of the agro-ecosystems’ (FAO, 2005).

Agroecology

‘The science and practice of applying ecological concepts, principles and knowledge (i.e., the interactions of, and explanations for, the diversity, abundance and activities of organisms) to the study, design and management of sustainable agroecosystems. It includes the roles of human beings as a central organism in agroecology by way of social and economic processes in farming systems. Agroecology examines the roles and interactions among all relevant biophysical, technical and socioeconomic components of farming systems and their surrounding landscapes’ (IPBES, 2019).

Agroforestry

Collective name for land-use systems and technologies where woody perennials (trees, shrubs, palms, bamboos, etc.) are deliberately used on the same land-management units as agricultural crops and/or animals, in some form of spatial arrangement or temporal sequence. In agroforestry systems there are both ecological and economical interactions between the different components. Agroforestry can also be defined as a dynamic, ecologically based, natural resource management system that, through the integration of trees on farms and in the agricultural landscape, diversifies and sustains production for increased social, economic and environmental benefits for land users at all levels (FAO, 2015a).

Air pollution

Degradation of air quality with negative effects on human health, the natural or built environment, due to the introduction by natural processes or human activity in the *atmosphere* of substances (gases, *aerosols*) which have a direct (primary pollutants) or indirect (secondary pollutants) harmful effect.

See also *Short-lived climate forcers (SLCF)*.

Albedo

The proportion of sunlight (solar radiation) reflected by a surface or object, often expressed as a percentage. Clouds, snow and ice usually have high albedo; soil surfaces cover the albedo range from high to low; vegetation in the dry season and/or in arid zones can have high albedo, whereas photosynthetically active vegetation and the ocean have low albedo. The Earth's planetary albedo changes mainly through varying cloudiness, snow, ice, leaf area and land cover changes.

Ambient persuasive technology

Technological systems and environments that are designed to change human cognitive processing, attitudes and behaviours without the need for the user's conscious attention.

Anomaly

The deviation of a variable from its value averaged over a *reference period*.

See also *Reference period*.

Anthromes

‘*Human systems*, with natural *ecosystems* embedded within them’ (Ellis and Ramankutty 2008). The anthrome classification system is based on human population density and *land use*, and comprises the following classes: dense settlements, villages, croplands, rangeland, forested (then broadened to seminatural) and wildlands (Ellis et al. 2010).

Anthropocene

A proposed new geological epoch resulting from significant human-driven changes to the structure and functioning of the Earth System, including the *climate system*. Originally proposed in the Earth System science community in 2000, the proposed new epoch is undergoing a formalization process within the geological community based on the stratigraphic *evidence* that human activities have changed the Earth System to the extent of forming geological deposits with a signature that is distinct from those of the *Holocene*, and which will remain in the geological record. Both the stratigraphic and Earth System approaches to defining the Anthropocene consider the mid-20th Century to be the most appropriate starting date, although others have been proposed and continue to be discussed. The Anthropocene concept has been taken up by a diversity of disciplines and the public to denote the substantive influence humans have had on the state, dynamics and future of the Earth System.

See also *Holocene*.

Anthropogenic

Resulting from or produced by human activities.

See also *Anthropogenic emissions*, and *Anthropogenic removals*.

Anthropogenic emissions

Emissions of *greenhouse gases* (GHGs), *precursors* of GHGs and *aerosols* caused by human activities. These activities include the burning of *fossil fuels*, *deforestation*, *land use and land use changes* (LULUC), livestock production, fertilisation, waste management, and industrial processes.

See also *Anthropogenic*, and *Anthropogenic removals*.

Anthropogenic removals

The withdrawal of *greenhouse gases* (GHGs) from the *atmosphere* as a result of deliberate human activities. These include enhancing biological *sinks* of CO₂ and using chemical engineering to achieve long term removal and storage. *Carbon capture and storage* (CCS) from industrial and energy-related sources, which alone does not remove CO₂ from the atmosphere, can help reduce atmospheric CO₂ if it is combined with *bioenergy production* (BECCS).

[Note: In the 2006 IPCC Guidelines for national GHG Inventories, which are used in reporting of emissions to the UNFCCC, ‘*anthropogenic*’ land-related GHG *fluxes* are defined as all those occurring on ‘*managed land*’, i.e. ‘where human interventions and practices have been applied to perform production, ecological or social functions’. However, some removals (e.g. removals associated with CO₂ fertilisation and N deposition) are not considered as ‘anthropogenic’, or are referred to as ‘indirect’ anthropogenic effects, in some of the scientific literature assessed in this report. As a consequence, the land-related net GHG emission estimates from global models included in this report are not necessarily directly comparable with LULUCF estimates in national GHG Inventories.]

See also *Anthropogenic emissions*, *Bioenergy with carbon dioxide capture and storage* (BECCS), *Carbon dioxide capture and storage* (CCS) and *Land use, Land-use change, and Forestry* (LULUCF).

Aridity

The state of a long-term climatic feature characterised by low average precipitation or available water in a region. Aridity generally arises from widespread persistent *atmospheric* subsidence or anticyclonic conditions, and from more localised subsidence in the lee side of mountains (adapted from Gbeckor-Kove, 1989; Türkeş, 1999).

Atmosphere

The gaseous envelope surrounding the Earth, divided into five layers — the *troposphere* which contains half of the Earth's atmosphere, the *stratosphere*, the mesosphere, the thermosphere, and the exosphere, which is the outer limit of the atmosphere. The dry atmosphere consists almost entirely of nitrogen (78.1% volume mixing ratio) and oxygen (20.9% volume mixing ratio), together with a number of trace gases, such as argon (0.93 % volume mixing ratio), helium and radiatively active *greenhouse gases (GHGs)* such as *carbon dioxide (CO₂)* (0.04% volume mixing ratio) and *ozone (O₃)*. In addition, the atmosphere contains the GHG water vapour (H₂O), whose amounts are highly variable but typically around 1% volume mixing ratio. The atmosphere also contains clouds and aerosols.

See also *Carbon dioxide (CO₂)*, *Ozone (O₃)*, *Troposphere*, *Stratosphere*, *Greenhouse gas (GHG)*, and *Hydrological cycle*.

Atmosphere-ocean general circulation model (AOGCM)

See *Climate model*.

Atmospheric boundary layer

The atmospheric layer adjacent to the Earth's surface that is affected by friction against that boundary surface, and possibly by transport of heat and other variables across that surface (AMS, 2000). The lowest 100 m of the boundary layer (about 10% of the boundary layer thickness), where mechanical generation of turbulence is dominant, is called the surface boundary layer or surface layer.

Attribution

See *Detection and attribution*.

Baseline scenario

In much of the literature the term is also synonymous with the term business-as-usual (BAU) *scenario*, although the term BAU has fallen out of favour because the idea of business as usual in century-long socio-economic *projections* is hard to fathom. In the context of *transformation pathways*, the term baseline scenarios refers to scenarios that are based on the assumption that no mitigation policies or measures will be implemented beyond those that are already in force and/or are legislated or planned to be adopted. Baseline scenarios are not intended to be predictions of the future, but rather counterfactual constructions that can serve to highlight the level of emissions that would occur without further policy effort. Typically, baseline scenarios are then compared to *mitigation scenarios* that are constructed to meet different goals for *greenhouse gas (GHG)* emissions, atmospheric concentrations or temperature change. The term baseline scenario is often used interchangeably with reference scenario and no policy scenario.

See also *Emission scenario*, and *Mitigation scenario*.

Biochar

Relatively stable, carbon-rich material produced by heating *biomass* in an oxygen-limited environment. Biochar is distinguished from charcoal by its application: biochar is used as a soil amendment with the intention to improve soil functions and to reduce *greenhouse gas (GHG)* emissions from biomass that would otherwise decompose rapidly (IBI, 2018).

Biodiversity

Biodiversity or biological diversity means the variability among living organisms from all sources including, among other things, terrestrial, marine and other aquatic *ecosystems* and the ecological complexes of which they are part; this includes diversity within species, between species and of ecosystems (UN, 1992).

See also *Ecosystem*, and *Ecosystem service*.

Bioenergy

Energy derived from any form of *biomass* or its metabolic by-products.

See also *Biomass* and *Biofuel*.

Bioenergy with carbon dioxide capture and storage (BECCS)

Carbon dioxide capture and storage (CCS) technology applied to a *bioenergy* facility. Note that depending on the total emissions of the BECCS supply chain, *carbon dioxide (CO₂)* can be removed from the *atmosphere*.

See also *Bioenergy*, and *Carbon dioxide capture and storage (CCS)*.

Biofuel

A fuel, generally in liquid form, produced from *biomass*. Biofuels include bioethanol from sugarcane, sugar beet or maize, and biodiesel from canola or soybeans.

See also *Biomass*, and *Bioenergy*.

Biogeochemical effects

Processes through which land affects *climate*, excluding *biophysical effects*. These processes include changes in net emissions of *carbon dioxide (CO₂)* towards the *atmosphere*, net emissions of *aerosols* (mineral and organic), ozone deposition on *ecosystems*, and net emissions of biogenic volatile organic compounds (BVOCs) and their subsequent changes in atmospheric chemistry.

See also *Biophysical effects*.

Biomass

Organic material excluding the material that is fossilised or embedded in geological formations. Biomass may refer to the mass of organic matter in a specific area (ISO, 2014).

See also *Bioenergy*, and *Biofuel*.

Traditional biomass

The combustion of wood, charcoal, agricultural residues and/or animal dung for cooking or heating in open fires or in inefficient stoves as is common in low-income countries.

Biome

‘Global-scale zones, generally defined by the type of plant life that they support in response to average rainfall and temperature patterns. For example, tundra, coral reefs or savannas’ (IPBES, 2019).

Biophysical effects

The range of physical processes through which *land* affects *climate*. These processes include changes in hydrology (e.g. water vapor fluxes at the land/atmosphere interface), heat exchanges via convective *fluxes* (latent and sensible), radiation (solar and infra-red, absorbed and emitted), and momentum (e.g. affecting wind speed).

Black carbon (BC)

A relatively pure form of carbon, also known as soot, arising from the incomplete combustion of *fossil fuels*, *biofuel*, and *biomass*. It stays in the *atmosphere* only for days or weeks. Black carbon is a climate forcing agent with strong warming effect, both in the atmosphere and when deposited on snow or ice.

See also *Atmosphere*, and *Aerosol*.

Blue carbon

All biologically-driven carbon *fluxes* and storage in marine systems that are amenable to management can be considered as blue carbon. Coastal blue carbon focuses on rooted vegetation in the coastal zone, such as tidal marshes, mangroves and seagrasses. These *ecosystems* have high carbon burial rates on a per unit area basis and accumulate carbon in their soils and sediments. They provide many non-climatic benefits and can contribute to ecosystem-based adaptation. If degraded or lost, coastal blue carbon ecosystems are likely to release most of their carbon back to the *atmosphere*. There is current debate regarding the application of the blue carbon concept to other coastal and non-coastal processes and ecosystems, including the open ocean.

See also *Ecosystem services*, and *Carbon sequestration*.

Business as usual (BAU)

See *Baseline scenario*.

Carbon budget

Refers to three concepts in the literature: (1) an assessment of *carbon cycle sources* and *sinks* on a global level, through the synthesis of *evidence* for *fossil-fuel* and cement emissions, *land-use change* emissions, ocean and land CO₂ sinks, and the resulting atmospheric *carbon dioxide (CO₂)* growth rate. This is referred to as the global carbon budget; (2) the estimated cumulative amount of global carbon dioxide emissions that that is estimated to limit global surface temperature to a given level above a *reference period*, taking into account global surface temperature contributions of other *greenhouse gases (GHGs)* and climate forcers; (3) the distribution of the carbon budget defined under (2) to the regional, national, or sub-national level based on considerations of *equity*, costs or efficiency.

See also *Remaining carbon budget*.

Carbon cycle

The flow of carbon (in various forms, e.g., as *carbon dioxide (CO₂)*, carbon in *biomass*, and carbon dissolved in the ocean as carbonate and bicarbonate) through the *atmosphere*, hydrosphere, terrestrial and marine biosphere and lithosphere. In this report, the reference unit for the global carbon cycle is GtCO₂ or GtC (one Gigatonne = 1 Gt = 10¹⁵ grams; 1GtC corresponds to 3.667 GtCO₂).

Carbon dioxide (CO₂)

A naturally occurring gas, CO₂ is also a by-product of burning *fossil fuels* (such as oil, gas and coal), of burning *biomass*, of *land use changes (LUC)* and of industrial processes (e.g., cement production). It is the principal *anthropogenic* greenhouse gas (*GHG*) that affects the Earth's radiative balance. It is the reference gas against which other GHGs are measured and therefore has a Global Warming Potential (GWP) of 1.

See also *Greenhouse gas (GHG)*, *Land use, Land-use change*, and *Global Warming Potential (GWP)*.

Carbon dioxide capture and storage (CCS)

A process in which a relatively pure stream of *carbon dioxide (CO₂)* from industrial and energy-related sources is separated (captured), conditioned, compressed and transported to a storage location for long-term isolation from the atmosphere. Sometimes referred to as Carbon Capture and Storage.

See also *Carbon dioxide capture and utilisation (CCU)*, *Bioenergy with carbon dioxide capture and storage (BECCS)*, and *Sequestration*.

Carbon dioxide capture and utilisation (CCU)

A process in which *carbon dioxide (CO₂)* is captured and then used to produce a new product. If the CO₂ is stored in a product for a *climate-relevant* time horizon, this is referred to as carbon dioxide capture, utilisation and storage (CCUS). Only then, and only combined with CO₂ recently removed from the *atmosphere*, can CCUS lead to carbon dioxide removal. CCU is sometimes referred to as Carbon dioxide capture and use.

See also *Carbon dioxide capture and storage (CCS)*.

Carbon dioxide capture, utilisation and storage (CCUS)

See *Carbon dioxide capture and utilisation (CCU)*.

Carbon dioxide removal (CDR)

Anthropogenic activities removing CO₂ from the *atmosphere* and durably storing it in geological, terrestrial, or ocean reservoirs, or in products. It includes existing and potential anthropogenic enhancement of biological or geochemical sinks and direct air capture and storage, but excludes natural CO₂ *uptake* not directly caused by human activities.

See also *Mitigation (of climate change)*, *Greenhouse gas removal (GGR)*, *Negative emission technologies*, and *Sink*.

Carbon intensity

The amount of emissions of *carbon dioxide (CO₂)* released per unit of another variable such as *gross domestic product (GDP)*, output energy use or transport.

Carbon price

The price for avoided or released *carbon dioxide (CO₂)* or *CO₂-equivalent emissions*. This may refer to the rate of a carbon tax, or the price of emission permits. In many models that are used to assess the economic costs of *mitigation*, carbon prices are used as a proxy to represent the level of effort in mitigation policies.

See also *Mitigation*.

Carbon sequestration

The process of storing carbon in a carbon *pool*.

See also *Blue carbon*, *Carbon dioxide capture and storage (CCS)*, *Uptake*, and *Sink*.

Carbon sink

See *Sink*.

Carbon stock

The quantity of carbon in a carbon *pool*.

See also *Pool*, *carbon and nitrogen*.

Citizen science

A voluntary participation of the public in the collection and/or processing of data as part of a scientific study (Silvertown, 2009).

Clean Development Mechanism (CDM)

A mechanism defined under Article 12 of the *Kyoto Protocol* through which investors (governments or companies) from developed (Annex B) countries may finance *greenhouse gas (GHG)* emission reduction or removal projects in developing countries (Non-Annex B), and receive Certified Emission Reduction Units (CERs) for doing so. The CERs can be credited towards the commitments of the respective developed countries. The CDM is intended to facilitate the two objectives of promoting *sustainable development (SD)* in developing countries and of helping *industrialised countries* to reach their emissions commitments in a cost-effective way.

Climate

Climate in a narrow sense is usually defined as the average weather, or more rigorously, as the statistical description in terms of the mean and variability of relevant quantities over a period of time ranging from months to thousands or millions of years. The classical period for averaging these variables is 30 years, as defined by the World Meteorological Organization. The relevant quantities are most often surface variables such as temperature, precipitation and wind. Climate in a wider sense is the state, including a statistical description, of the *climate system*.

Climate change

A change in the state of the *climate* that can be identified (e.g., by using statistical tests) by changes in the mean and/or the variability of its properties and that persists for an extended period, typically decades or longer. Climate change may be due to natural internal processes or external *forcings* such as modulations of the solar cycles, volcanic eruptions and persistent *anthropogenic* changes in the composition of the *atmosphere* or in *land use*. Note that the *United Nations Framework Convention on Climate Change (UNFCCC)*, in its Article 1, defines climate change as: 'a change of climate which is attributed directly or indirectly to human activity that alters the composition of the global atmosphere and which is in addition to natural climate variability observed over comparable time periods'. The UNFCCC thus makes a distinction between climate change attributable to human activities altering the atmospheric composition and climate variability attributable to natural causes.

See also *Climate variability*, *Global warming*, *Ocean acidification*, and *Detection and attribution*.

Climate extreme (extreme weather or climate event)

The occurrence of a value of a weather or *climate* variable above (or below) a threshold value near the upper (or lower) ends of the range of observed values of the variable. For simplicity, both extreme weather events and extreme climate events are referred to collectively as ‘climate extremes.’

See also *Extreme weather event*.

Climate feedback

An interaction in which a perturbation in one *climate* quantity causes a change in a second and the change in the second quantity ultimately leads to an additional change in the first. A negative feedback is one in which the initial perturbation is weakened by the changes it causes; a positive feedback is one in which the initial perturbation is enhanced. The initial perturbation can either be externally forced or arise as part of internal variability.

Climate governance

See *Governance*.

Climate model

A numerical representation of the *climate system* based on the physical, chemical and biological properties of its components, their interactions and *feedback* processes and accounting for some of its known properties. The climate system can be represented by models of varying complexity; that is, for any one component or combination of components a spectrum or hierarchy of models can be identified, differing in such aspects as the number of spatial dimensions, the extent to which physical, chemical or biological processes are explicitly represented, or the level at which empirical parametrizations are involved. There is an evolution towards more complex models with interactive chemistry and biology. *Climate models* are applied as a research tool to study and simulate the *climate* and for operational purposes, including monthly, seasonal and interannual climate predictions.

See also *Earth system model (ESM)*.

Climate projection

Simulated response of the *climate system* to a *scenario* of future emissions or concentrations of *greenhouse gases (GHGs)* and *aerosols*, and changes in land use, generally derived using *climate models*. Climate projections are distinguished from climate predictions by their dependence on the emission/concentration/*radiative forcing* scenario used, which is in turn based on assumptions concerning, for example, future socioeconomic and technological developments that may or may not be realised.

Climate-resilient development pathways (CRDPs)

Trajectories that strengthen *sustainable development* and efforts to eradicate *poverty* and reduce inequalities while promoting *fair* and cross-scalar *adaptation* to and *resilience* in a changing *climate*. They raise the *ethics*, *equity*, and *feasibility* aspects of the deep *societal transformation* needed to drastically reduce emissions to limit *global warming* (e.g., to 2°C) and achieve desirable and liveable futures and *well-being* for all.

Climate-resilient pathways

Iterative processes for managing change within complex systems in order to reduce disruptions and enhance opportunities associated with *climate change*.

See also *Climate-resilient development pathways (CRDPs)*, *Development pathways*, *Pathways*, and *Transformation pathways*.

Climate sensitivity

The change in the annual *global mean surface temperature* in response to a change in the atmospheric *carbon dioxide (CO₂)* concentration or other radiative forcing.

Equilibrium climate sensitivity

The equilibrium (steady state) change in the globally-averaged near-surface temperature following a doubling of the atmospheric *carbon dioxide (CO₂)* concentration from pre-industrial conditions. The equilibrium climate sensitivity is often estimated through experiments in atmosphere-ocean general circulation model (AOGCMs) where CO₂ levels are either quadrupled or doubled from *pre-industrial* levels and which are integrated for 100-200 years. A related quantity, the climate feedback parameter (units: W m⁻² °C⁻¹) refers to the top of atmosphere budget change per degree of globally-averaged near-surface temperature change.

See also *Climate model*, and *Global mean surface temperature (GMST)*.

Effective climate sensitivity

An estimate of the *global mean surface temperature* response to a doubling of the atmospheric *carbon dioxide (CO₂)* concentration that is evaluated from model output or observations for evolving non-equilibrium conditions. It is a measure of the strengths of the *climate feedbacks* at a particular time and may vary with *forcing* history and *climate* state, and therefore may differ from *equilibrium climate sensitivity*.

Transient climate response

The change in the *global mean surface temperature*, averaged over a 20-year period, centred at the time of atmospheric *carbon dioxide (CO₂)* doubling, in a *climate model* simulation in which CO₂ increases at 1% yr⁻¹ from *pre-industrial*. It is a measure of the strength of *climate feedbacks* and the timescale of ocean heat uptake.

Climate services

Information and products that enhance users' knowledge and understanding about the *impacts* of *climate change* and/or *climate variability* so as to aid decision-making of individuals and organizations and enable preparedness and early climate change action. Such services involve high-quality data from national and international databases on temperature, rainfall, wind, soil moisture and ocean conditions, as well as maps, *risk* and *vulnerability* analyses, assessments, and long-term *projections* and *scenarios*. Depending on the user's needs, these data and information products may be combined with non-meteorological data, such as agricultural production, health trends, population distributions in high-risk areas, road and infrastructure maps for the delivery of goods, and other socio-economic variables (WMO, 2019).

Climate-smart agriculture (CSA)

An approach to agriculture that aims to transform and reorient agricultural systems to effectively support development and ensure *food security* in a changing climate by: sustainably increasing agricultural productivity and incomes; adapting and building resilience to climate change; and reducing and/or removing *greenhouse gas* emissions, where possible (FAO, 2018).

Climate system

The system consisting of five major components: the *atmosphere*, the hydrosphere, the cryosphere, the lithosphere and the biosphere and the interactions between them. The climate system evolves in time under the influence of its own internal dynamics and because of external *forcings* such as volcanic eruptions, solar variations, orbital forcing, and *anthropogenic* forcings such as the changing composition of the atmosphere and *land-use change*.

Climate target

A temperature limit, concentration level, or emissions reduction goal used towards the aim of avoiding dangerous *anthropogenic* interference with the *climate system*. For example, national climate targets may aim to reduce *greenhouse gas* emissions by a certain amount over a given time horizon, for example those under the *Kyoto Protocol*.

Climate variability

Variations in the mean state and other statistics (such as standard deviations, the occurrence of extremes, etc.) of the *climate* on all spatial and temporal scales beyond that of individual weather events. Variability may be due to natural internal processes within the *climate system* (internal variability), or to variations in natural or *anthropogenic* external *forcing* (external variability).

See also *Climate change*.

CO₂ equivalent (CO₂-eq) emission

The amount of *carbon dioxide* (CO₂) emission that would cause the same integrated *radiative forcing* or temperature change, over a given time horizon, as an emitted amount of a *greenhouse gas* (GHG) or a mixture of GHGs. There are a number of ways to compute such equivalent emissions and choose appropriate time horizons. Most typically, the CO₂-equivalent emission is obtained by multiplying the emission of a GHG by its Global Warming Potential (GWP) for a 100 year time horizon. For a mix of GHGs it is obtained by summing the CO₂-equivalent emissions of each gas. CO₂-equivalent emission is a common scale for comparing emissions of different GHGs but does not imply equivalence of the corresponding climate change responses. There is generally no connection between CO₂-equivalent emissions and resulting CO₂-equivalent concentrations.

Co-benefits

The positive effects that a policy or measure aimed at one objective might have on other objectives, thereby increasing the total benefits for society or the environment. Co-benefits are often subject to uncertainty and depend on local circumstances and implementation practices, among other factors. Co-benefits are also referred to as ancillary benefits.

See also *Adverse side-effects*, and *Risk*.

Collective action

A number of people working together voluntarily to achieve some common objective (Meinzen-Dick and Di Gregorio, 2004).

Conference of the Parties (COP)

The supreme body of UN conventions, such as the *United Nations Framework Convention on Climate Change* (UNFCCC), comprising parties with a right to vote that have ratified or acceded to the convention.

See also *United Nations Convention to Combat Desertification* (UNCCD).

Confidence

The robustness of a finding based on the type, amount, quality and consistency of *evidence* (e.g., mechanistic understanding, theory, data, models, expert judgment) and on the degree of *agreement* across multiple lines of evidence. In this report, confidence is expressed qualitatively (Mastrandrea et al., 2010).

See also *Likelihood*, and *Uncertainty*.

Convection

Vertical motion driven by buoyancy forces arising from static instability, usually caused by near-surface cooling or increases in salinity in the case of the ocean and near-surface warming or cloud-top radiative cooling in the case of the *atmosphere*. In the atmosphere, convection gives rise to cumulus clouds and precipitation and is effective at both scavenging and vertically transporting chemical species. In the ocean, convection can carry surface waters to deep within the ocean.

Coping capacity

The ability of people, *institutions*, organizations, and systems, using available skills, values, beliefs, resources, and opportunities, to address, manage, and overcome adverse conditions in the short to medium term. (UNISDR, 2009; IPCC, 2012a)

See also *Resilience*.

Cost-benefit analysis

Monetary assessment of all negative and positive impacts associated with a given action. Cost-benefit analysis enables comparison of different interventions, investments or strategies and reveal how a given investment or policy effort pays off for a particular person, company or country. Cost-benefit analyses representing society's point of view are important for *climate change* decision making, but there are difficulties in aggregating costs and benefits across different actors and across timescales.

See also *Discounting*.

Cost-effectiveness

A measure of the cost at which a policy goal or outcome is achieved. The lower the cost the greater the cost-effectiveness.

See also *Integrated models*.

Coupled Model Intercomparison Project (CMIP)

A climate modelling activity from the World Climate Research Programme (WCRP) which coordinates and archives *climate model* simulations based on shared model inputs by modelling groups from around the world. The CMIP3 multi-model data set includes *projections* using SRES scenarios. The CMIP5 data set includes projections using the *Representative Concentration Pathways (RCPs)*. The CMIP6 phase involves a suite of common model experiments as well as an ensemble of CMIP-endorsed model intercomparison projects (MIPs).

Cumulative emissions

The total amount of emissions released over a specified period of time.

See also *Carbon budget*.

Decarbonisation

Process by which countries, individuals or other entities aim to achieve zero fossil carbon existence. Typically refers to a reduction of the carbon emissions associated with electricity, industry and transport.

Decoupling

Decoupling (in relation to *climate change*) is where economic growth is no longer strongly associated with consumption of *fossil fuels*. Relative decoupling is where both grow but at different rates. Absolute decoupling is where economic growth happens but fossil fuels decline.

Deforestation

Conversion of *forest* to non-forest.

[Note: For a discussion of the term forest and related terms such as *afforestation*, *reforestation* and *deforestation* in the context of reporting and accounting Article 3.3 and 3.4 activities under the Kyoto Protocol, see 2013 Revised Supplementary Methods and Good Practice Guidance Arising from the Kyoto Protocol.]

See also *Reducing Emissions from Deforestation and Forest Degradation (REDD+)*.

Deliberative governance

See *Governance*.

Demand and supply-side measures*Demand-side measures*

Policies and programmes for influencing the demand for goods and/or services. In the energy sector, demand-side management aims at reducing the demand for electricity and other forms of energy required to deliver energy services.

Supply-side measures

Policies and programmes for influencing how a certain demand for goods and/or services is met. In the energy sector, for example, supply-side *mitigation measures* aim at reducing the amount of *greenhouse gas (GHG)* emissions emitted per unit of energy produced.

See also *Mitigation measures*.

Demand-side measures

See *Demand and supply-side measures*.

Desertification

Land degradation in arid, semi-arid, and dry sub-humid areas resulting from many factors, including climatic variations and human activities (UNCCD, 1994).

Detection

See *Detection and attribution*.

Detection and attribution

Detection of change is defined as the process of demonstrating that *climate* or a system affected by climate has changed in some defined statistical sense, without providing a reason for that change. An identified change is detected in observations if its *likelihood* of occurrence by chance due to internal variability alone is determined to be small, for example, <10%. Attribution is defined as the process of evaluating the relative contributions of multiple causal factors to a change or event with a formal assessment of *confidence*.

Development pathways

See *Pathways*.

Diet

‘The kinds of food that follow a particular pattern that a person or community eats’ (FAO, 2014).

See also *Dietary patterns*.

Dietary patterns

The quantities, proportions, variety or combinations of different foods and beverages in diets, and the frequency with which they are habitually consumed (Dietary Guidelines Advisory Committee, 2015).

See also *Diet*.

Dietary and nutrition transitions

Modernization, urbanisation, economic development, and increased wealth lead to predictable shifts in diet, referred to as ‘nutrition transitions’ (Misra and Khurana, 2008; Popkin, 2006). Over historical time there have been a number of dietary *transitions* but in recent decades the prime transition has been associated with changes from subsistence towards eating diets rich in calories and relatively poor in nutrition (the ‘westernised diet’) that are obesogenic. From a public health perspective, a new dietary transition is in focus, from the obesogenic diet to one promoting health.

Disaster

A ‘serious disruption of the functioning of a community or a society at any scale due to hazardous events interacting with conditions of exposure, vulnerability and capacity, leading to one or more of the following: human, material, economic and environmental losses and impacts’ (UNISDR, 2017).

See also *Exposure, Risk, Vulnerability and Hazard*.

Disaster risk management (DRM)

Processes for designing, implementing, and evaluating strategies, policies, and measures to improve the understanding of current and future *disaster* risk, foster disaster risk reduction and transfer, and promote continuous improvement in disaster preparedness, prevention and protection, response, and recovery practices, with the explicit purpose of increasing human security, *well-being*, quality of life, and *sustainable development* (UNISDR, 2017).

Discounting

A mathematical operation that aims to make monetary (or other) amounts received or expended at different times (years) comparable across time. The discounter uses a fixed or possibly time-varying discount rate from year to year that makes future value worth less today (if the discount rate is positive). The choice of discount rate(s) is debated as it is a judgement based on hidden and/or explicit values.

Discount rate

See *Discounting*.

(Internal) Displacement

The forced movement of people within the country they live in. Internally displaced persons (IDPs) are 'Persons or groups of persons who have been forced or obliged to flee or to leave their homes or places of habitual residence, in particular as a result of or in order to avoid the effects of armed conflict, situations of generalized violence, violations of human rights or natural or human-made *disasters*, and who have not crossed an internationally recognized State border' (UN, 1998).

See also *Migration*.

Displacement

In land system science, displacement denotes the increasing spatial separation between the location of agricultural and forestry production and the place of consumption of these products, as it occurs with trade. Displacement disconnects spatially environmental impacts from their socioeconomic drivers.

Downscaling

Method that derives local- to regional-scale (up to 100 km) information from larger-scale models or data analyses. Two main methods exist: dynamical downscaling and empirical/statistical downscaling. The dynamical method uses the output of regional *climate models*, global models with variable spatial resolution, or high-resolution global models. The empirical/statistical methods [are based on observations and] develop statistical relationships that link the large-scale atmospheric variables with local/ regional climate variables. In all cases, the quality of the driving model remains an important limitation on quality of the downscaled information. The two methods can be combined, e.g., applying empirical/statistical downscaling to the output of a regional climate model, consisting of a dynamical downscaling of a global climate model.'

Drainage

'Artificial lowering of the soil water table' (IPCC, 2013).

See also *Rewetting*.

Drought

A period of abnormally dry weather long enough to cause a serious hydrological imbalance. Drought is a relative term, therefore any discussion in terms of precipitation deficit must refer to the particular precipitation-related activity that is under discussion. For example, shortage of precipitation during the growing season impinges on crop production or *ecosystem* function in general (due to *soil moisture* drought, also termed agricultural drought), and during the *runoff* and percolation season primarily affects water supplies (hydrological drought). Storage changes in soil moisture and groundwater are also affected by increases in actual *evapotranspiration* in addition to reductions in

precipitation. A period with an abnormal precipitation deficit is defined as a meteorological drought.

Megadrought

A very lengthy and pervasive drought, lasting much longer than normal, usually a decade or more.

Early warning systems (EWS)

The set of technical, financial and *institutional capacities* needed to generate and disseminate timely and meaningful warning information to enable individuals, communities and organizations threatened by a *hazard* to prepare to act promptly and appropriately to reduce the possibility of harm or loss. Dependent upon context, EWS may draw upon scientific and/or *Indigenous knowledge*. EWS are also considered for ecological applications e.g., conservation, where the organisation itself is not threatened by hazard but the *ecosystem* under conservation is (an example is coral bleaching alerts), in agriculture (for example, warnings of ground frost, hailstorms) and in fisheries (storm and tsunami warnings) (UNISDR, 2009; IPCC, 2012a).

Earth system feedbacks

See *Climate feedback*.

Earth system model (ESM)

A coupled atmosphere–ocean general circulation model in which a representation of the *carbon cycle* is included, allowing for interactive calculation of atmospheric *carbon dioxide* (CO_2) or compatible emissions. Additional components (e.g., atmospheric chemistry, *ice sheets*, dynamic vegetation, nitrogen cycle, but also urban or crop models) may be included.

See also *Climate model*.

Ecological cascade

A series of secondary extinctions as a result of the extinction of a key species within an *ecosystem* (Soulé, 2010).

Ecosystem

A functional unit consisting of living organisms, their non-living environment and the interactions within and between them. The components included in a given ecosystem and its spatial boundaries depend on the purpose for which the ecosystem is defined: in some cases they are relatively sharp, while in others they are diffuse. Ecosystem boundaries can change over time. Ecosystems are nested within other ecosystems and their scale can range from very small to the entire biosphere. In the current era, most ecosystems either contain people as key organisms, or are influenced by the effects of human activities in their environment.

See also *Ecosystem services*.

Ecosystem services

Ecological processes or functions having monetary or non-monetary value to individuals or society at large. These are frequently classified as (1) supporting services such as productivity or *biodiversity* maintenance, (2) provisioning services such as food or fibre, (3) regulating services such as climate regulation or *carbon sequestration*, and (4) cultural services such as tourism or spiritual and aesthetic appreciation.

See also *Ecosystem*.

Effective climate sensitivity

See *Climate sensitivity*.

Effective radiative forcing

See *Radiative forcing*.

El Niño-Southern Oscillation (ENSO)

The term El Niño was initially used to describe a warm-water current that periodically flows along the coast of Ecuador and Peru, disrupting the local fishery. It has since become identified with warming of the tropical Pacific Ocean east of the dateline. This oceanic event is associated with a fluctuation of a global-scale tropical and subtropical surface pressure pattern called the Southern Oscillation. This coupled atmosphere–ocean phenomenon, with preferred time scales of two to about seven years, is known as the El Niño-Southern Oscillation (ENSO). It is often measured by the surface pressure anomaly difference between Tahiti and Darwin and/or the *sea surface temperatures* in the central and eastern equatorial Pacific. During an ENSO event, the prevailing trade winds weaken, reducing upwelling and altering ocean currents such that the sea surface temperatures warm, further weakening the trade winds. This phenomenon has a great impact on the wind, sea surface temperature and precipitation patterns in the tropical Pacific. It has climatic effects throughout the Pacific region and in many other parts of the world, through global *teleconnections*. The cold phase of ENSO is called La Niña.

Embodied (embedded) [emissions, water, land]

The total emissions [water use, *land use*] generated [used] in the production of goods and services regardless of the location and timing of those emissions [water use, *land use*] in the production process. This includes emissions [water use, *land use*] within the country used to produce goods or services for the country’s own use, but also includes the emissions [water use, *land use*] related to the production of such goods or services in other countries that are then consumed in another country through imports. Such emissions [water, *land*] are termed ‘embodied’ or ‘embedded’ emissions, or in some cases (particularly with water) as ‘virtual water use’ (David and Caldeira, 2010; Allan, 2005; MacDonald et al., 2015).

Emission scenario

A plausible representation of the future development of emissions of substances that are radiatively active (e.g., *greenhouse gases (GHGs)*, *aerosols*) based on a coherent and internally consistent set of assumptions about driving forces (such as demographic and socio-economic development, technological change, energy and *land use*) and their key relationships. Concentration *scenarios*, derived from emission scenarios, are often used as input to a *climate model* to compute *climate projections*.

See also *Baseline scenario*, *Mitigation scenario*, *Representative Concentration Pathways (RCPs) (under Pathways)*, *Shared socio-economic pathways (SSPs) (under Pathways)*, *Scenario*, *Socio-economic scenario*, and *Transformation pathway*.

Emission trajectories

A *projected* development in time of the emission of a *greenhouse gas (GHG)* or group of GHGs, *aerosols*, and GHG *precursors*.

See also *Pathways*.

Energy access

Access to clean, reliable and affordable energy services for cooking and heating, lighting, communications, and productive uses (with special reference to *Sustainable Development Goal 7*) (AGECC, 2010).

See also *Traditional biomass*.

Enabling conditions (for adaptation and mitigation options)

Conditions that affect the *feasibility* of *adaptation* and *mitigation* options, and can accelerate and scale-up systemic transitions that would limit temperature increase and enhance capacities of systems and societies to adapt to the associated *climate change*, while achieving *sustainable development*, eradicating *poverty* and reducing *inequalities*. Enabling conditions include finance, technological innovation, strengthening policy instruments, *institutional capacity*, *multi-level governance*, and changes in *human behaviour* and lifestyles. They also include inclusive processes, attention to power asymmetries and unequal opportunities for development and reconsideration of values.

Energy efficiency

The ratio of output or useful energy or energy services or other useful physical outputs obtained from a system, conversion process, transmission or storage activity to the input of energy (measured as kWh kWh⁻¹, tonnes kWh⁻¹ or any other physical measure of useful output like tonne-km transported). Energy efficiency is often described by energy intensity. In economics, energy intensity describes the ratio of economic output to energy input. Most commonly energy efficiency is measured as input energy over a physical or economic unit, i.e. kWh USD⁻¹ (energy intensity), kWh tonne⁻¹. For buildings, it is often measured as kWh m⁻², and for vehicles as km liter⁻¹ or liter km⁻¹. Very often in policy ‘energy efficiency’ is intended as the measures to reduce energy demand through technological options such as insulating buildings, more efficient appliances, efficient lighting, efficient vehicles, etc.

Energy security

The goal of a given country, or the global community as a whole, to maintain an adequate, stable and predictable energy supply. Measures encompass safeguarding the sufficiency of energy resources to meet national energy demand at competitive and stable prices and the resilience of the energy supply; enabling development and deployment of technologies; building sufficient infrastructure to generate, store and transmit energy supplies and ensuring enforceable contracts of delivery.

Enhanced weathering

Enhancing the removal of *carbon dioxide (CO₂)* from the *atmosphere* through dissolution of silicate and carbonate rocks by grinding these minerals to small particles and actively applying them to soils, coasts or oceans.

(Model) Ensemble

A group of parallel model simulations characterising historical climate conditions, *climate* predictions, or *climate projections*. Variation of the results across the ensemble members may give an estimate of modelling-based *uncertainty*. Ensembles made with the same model but different initial conditions only characterise the uncertainty associated with internal *climate variability*, whereas multi-model ensembles including simulations by several models also include the impact of model differences. Perturbed parameter ensembles, in which model parameters are varied in a systematic manner, aim to assess the uncertainty resulting from internal model specifications within a single

model. Remaining sources of uncertainty unaddressed with model ensembles are related to systematic model errors or biases, which may be assessed from systematic comparisons of model simulations with observations wherever available.

See also *Climate projection*.

Equality

A principle that ascribes equal worth to all human beings, including equal opportunities, rights, and obligations, irrespective of origins.

Inequality

Uneven opportunities and social positions, and processes of discrimination within a group or society, based on gender, class, ethnicity, age, and (dis)ability, often produced by uneven development. Income inequality refers to gaps between highest and lowest income earners within a country and between countries.

See also *Equity and Fairness*.

Equilibrium climate sensitivity

See *Climate sensitivity*.

Equity

The principle of being fair and impartial, and a basis for understanding how the *impacts* and responses to *climate change*, including costs and benefits, are distributed in and by society in more or less equal ways. It is often aligned with ideas of *equality*, fairness and justice and applied with respect to *equity* in the responsibility for, and distribution of, climate impacts and policies across society, generations, and gender, and in the sense of who participates and controls the processes of decision making.

Distributive equity

Equity in the consequences, outcomes, costs and benefits of actions or policies. In the case of *climate change* or climate policies for different people, places and countries, including equity aspects of sharing burdens and benefits for *mitigation* and *adaptation*.

Gender equity

Equity between women and men with regard to their rights, resources and opportunities. In the case of *climate change*, gender equity recognises that women are often more vulnerable to the *impacts* of climate change and may be disadvantaged in the process and outcomes of climate policy.

Inter-generational equity

Equity between generations. In the context of *climate change*, inter-generational equity acknowledges that the effects of past and present emissions, *vulnerabilities* and policies impose costs and benefits for people in the future and of different age groups.

Procedural equity

Equity in the process of decision making including recognition and inclusiveness in participation, equal representation, bargaining power, voice and equitable access to knowledge and resources to participate.

See also *Equality and Fairness*.

Evaporation

The physical process by which a liquid (e.g., water) becomes a gas (e.g., water vapour).

Evapotranspiration

The combined processes through which water is transferred to the *atmosphere* from open water and ice surfaces, bare soil, and vegetation that make up the Earth's surface.

Potential Evapotranspiration

The potential rate of water loss without any limits imposed by the water supply.

See also *Evaporation*.

Evidence

Data and information used in the scientific process to establish findings. In this report, the degree of evidence reflects the amount, quality, and consistency of scientific/technical information on which the Lead Authors base their findings.

See also *Agreement, Confidence, Likelihood, and Uncertainty*.

Exposure

The presence of people; *livelihoods*; species or *ecosystems*; environmental functions, services, and resources; infrastructure; or economic, social, or cultural assets in places and settings that could be adversely affected.

See also *Hazard, Risk, and Vulnerability*.

Extratropical Cyclone

Any cyclonic-scale storm that is not a *tropical cyclone*. Usually refers to a middle- or high-latitude migratory storm system formed in regions of large horizontal temperature variations. Sometimes called extratropical storm or extratropical low.

See also *Tropical cyclone*.

Extreme weather or climate event

See *Climate extreme (extreme weather or climate event)*.

Extreme weather event

An event that is rare at a particular place and time of year. Definitions of 'rare' vary, but an extreme weather event would normally be as rare as or rarer than the 10th or 90th percentile of a probability density function estimated from observations. By definition, the characteristics of what is called extreme weather may vary from place to place in an absolute sense. When a pattern of extreme weather persists for some time, such as a season, it may be classed as an extreme climate event, especially if it yields an average or total that is itself extreme (e.g., *drought* or heavy rainfall over a season).

See also *Heat wave, and Climate extreme (extreme weather or climate event)*.

Fairness

Impartial and just treatment without favouritism or discrimination in which each person is considered of equal worth with equal opportunity.

See also *Equity*, and *Equality*.

Feasibility

The degree to which climate goals and response options are considered possible and/or desirable. Feasibility depends on geophysical, ecological, technological, economic, social and *institutional* conditions for change. Conditions underpinning feasibility are dynamic, spatially variable, and may vary between different groups.

See also *Enabling conditions*.

Feedback

See *Climate feedback*.

Flexible governance

See *Governance*.

Flood

The overflowing of the normal confines of a stream or other body of water, or the accumulation of water over areas that are not normally submerged. Floods include river (fluvial) floods, flash floods, urban floods, rain (pluvial) floods, sewer floods, coastal floods, and glacial lake outburst floods.

Flux

A movement (a flow) of matter (e.g., water vapor, particles), heat or energy from one place to another, or from one medium (e.g., land surface) to another (e.g., atmosphere).

Food loss and waste

‘The decrease in quantity or quality of food’. Food waste is part of food loss and refers to discarding or alternative (non-food) use of food that is safe and nutritious for human consumption along the entire food supply chain, from primary production to end household consumer level. Food waste is recognised as a distinct part of food loss because the drivers that generate it and the solutions to it are different from those of food losses (FAO, 2015b).

Food security

A situation that exists when all people, at all times, have physical, social and economic access to sufficient, safe and nutritious food that meets their dietary needs and food preferences for an active and healthy life (FAO, 2001).

[Note: Whilst the term ‘food security’ explicitly includes nutrition within it ‘dietary needs ...for an active and healthy life’, in the past the term has sometimes privileged the supply of calories (energy), especially to the hungry. Thus, the term ‘food and nutrition security’ is often used (with the same definition as food security) to emphasise that the term food covers both energy and nutrition (FAO, 2009).]

Food system

All the elements (environment, people, inputs, processes, infrastructures, institutions, etc.) and activities that relate to the production, processing, distribution, preparation and consumption of food, and the output of these activities, including socio-economic and environmental outcomes (HLPE, 2017).

[Note: Whilst there is a global food system (encompassing the totality of global production and consumption), each location's food system is unique, being defined by that place's mix of food produced locally, nationally, regionally or globally.]

Forcing

See *Radiative forcing*.

Forest

A vegetation type dominated by trees. Many definitions of the term forest are in use throughout the world, reflecting wide differences in biogeophysical conditions, social structure and economics.

[Note: For a discussion of the term forest in the context of National GHG inventories, see the 2006 IPCC Guidelines for National GHG Inventories and information provided by the United Nations Framework Convention on Climate Change (UNFCCC, 2019).]

See also *Afforestation*, *Deforestation*, and *Reforestation*.

Fossil fuels

Carbon-based fuels from fossil hydrocarbon deposits, including coal, oil, and natural gas.

Framework Convention on Climate Change

See *United Nations Framework Convention on Climate Change (UNFCCC)*.

Gender equity

See *Equity*.

Glacier

A perennial mass of ice, and possibly firn and snow, originating on the *land* surface by the recrystallisation of snow and showing *evidence* of past or present flow. A glacier typically gains mass by accumulation of snow, and loses mass by melting and ice discharge into the sea or a lake if the glacier terminates in a body of water. Land ice masses of continental size (>50 000 km²) are referred to as *ice sheets*.

Global climate model (also referred to as general circulation model, both abbreviated as GCM)

See *Climate model*.

Global mean surface temperature (GMST)

Estimated global average of near-surface air temperatures over land and sea-ice, and *sea surface temperatures* over ice-free ocean regions, with changes normally expressed as departures from a value over a specified *reference period*. When estimating changes in GMST, near-surface air temperature over both land and oceans are also used.

See also *Global mean surface air temperature (GSAT)*, *Land surface air temperature*, and *Sea surface temperature (SST)*.

Global mean surface air temperature (GSAT)

Global average of near-surface air temperatures over land and oceans. Changes in GSAT are often used as a measure of global temperature change in *climate models* but are not observed directly.

See also *Global mean surface temperature (GMST)*, and *Land surface air temperature*.

Global warming

An increase in *global mean surface temperature (GMST)* averaged over a 30-year period, or the 30-year period centred on a particular year or decade, expressed relative to *pre-industrial* levels unless otherwise specified. For 30-year periods that span past and future years, the current multi-decadal warming trend is assumed to continue.

See also *Climate change*, and *Climate variability*.

Governance

A comprehensive and inclusive concept of the full range of means for deciding, managing, implementing and monitoring policies and measures. Whereas government is defined strictly in terms of the nation-state, the more inclusive concept of governance recognises the contributions of various levels of government (global, international, regional, sub-national and local) and the contributing roles of the private sector, of nongovernmental actors, and of civil society to addressing the many types of issues facing the global community, and the local context where the effectiveness of policies and measures are determined.

Adaptive governance

An emerging term in the literature for the evolution of formal and informal *institutions* of governance that prioritise planning, implementation and evaluation of policy through iterative *social learning*; in the context of *climate change*, governance facilitating social learning to steer the use and protection of natural resources, and *ecosystem services*, particularly in situations of complexity and *uncertainty*.

Climate governance

Purposeful mechanisms and measures aimed at steering social systems towards preventing, mitigating, or adapting to the risks posed by climate change (Jagers and Striiple, 2003).

Deliberative governance

Involves decision making through inclusive public conversation which allows opportunity for developing policy options through public discussion rather than collating individual preferences through voting or referenda (although the latter governance mechanisms can also be proceeded and legitimated by public deliberation processes).

Flexible governance

Strategies of governance at various levels, which prioritise the use of *social learning* and rapid feedback mechanisms in planning and policy making, often through incremental, experimental and iterative management processes.

Governance capacity

The ability of governance *institutions*, leaders, and non-state and civil society to plan, co-ordinate, fund, implement, evaluate and adjust policies and measures over the short, medium and long term, adjusting for *uncertainty*, rapid change and wide ranging impacts and multiple actors and demands.

Multi-level governance

Negotiated, non-hierarchical exchanges between *institutions* at the transnational, national, regional and local levels.

Participatory governance

A governance system that enables direct public engagement in decision-making using a variety of techniques for example, referenda, community deliberation, citizen juries or participatory budgeting. The approach can be applied in formal and informal *institutional* contexts from national to local, but is usually associated with devolved decision making (Fung and Wright, 2003; Sarmiento and Tilly, 2018).

Governance capacity

See *Governance*.

Grazing land

The sum of rangelands and *pastures* not considered as cropland, and subject to livestock grazing or hay production. It includes a wide range of *ecosystems*, e.g. systems with vegetation that fall below the threshold used in the *forest* land category, silvo-pastoral systems, as well as natural, managed grasslands and semideserts.

Green infrastructure

The interconnected set of natural and constructed ecological systems, green spaces and other landscape features. It includes planted and indigenous trees, wetlands, parks, green open spaces and original grassland and woodlands, as well as possible building and street level design interventions that incorporate vegetation. Green infrastructure provides services and functions in the same way as conventional infrastructure (Culwick and Bobbins, 2016).

Greenhouse gas (GHG)

Gaseous constituents of the *atmosphere*, both natural and *anthropogenic*, that absorb and emit radiation at specific wavelengths within the spectrum of terrestrial radiation emitted by the Earth's surface, the atmosphere itself, and by clouds. This property causes the greenhouse effect. Water vapour (H₂O), *carbon dioxide* (CO₂), *nitrous oxide* (N₂O), *methane* (CH₄) and *ozone* (O₃) are the primary GHGs in the Earth's atmosphere. Moreover, there are a number of entirely human-made GHGs in the atmosphere, such as the *halocarbons* and other chlorine- and bromine-containing substances, dealt with under the Montreal Protocol. Beside CO₂, N₂O and CH₄, the *Kyoto Protocol* deals with the GHGs sulphur hexafluoride (SF₆), hydrofluorocarbons (HFCs) and perfluorocarbons (PFCs).

Greenhouse gas removal (GGR)

Withdrawal of a *greenhouse gas* (GHG) and/or a *precursor* from the *atmosphere* by a *sink*.

See also *Carbon dioxide removal* (CDR), and *Negative emissions*.

Gross domestic product (GDP)

The sum of gross value added, at purchasers' prices, by all resident and non-resident producers in the economy, plus any taxes and minus any subsidies not included in the value of the products in a country or a geographic region for a given period, normally one year. GDP is calculated without deducting for depreciation of fabricated assets or depletion and degradation of natural resources.

Halocarbons

A collective term for the group of partially halogenated organic species, which includes the chlorofluorocarbons (CFCs), hydrochlorofluorocarbons (HCFCs), hydrofluorocarbons (HFCs), halons, methyl chloride and methyl bromide. Many of the halocarbons have large Global Warming Potentials. The chlorine and bromine-containing halocarbons are also involved in the depletion of the ozone layer.

Hazard

The potential occurrence of a natural or human-induced physical event or trend that may cause loss of life, injury, or other health impacts, as well as damage and loss to property, infrastructure, *livelihoods*, service provision, *ecosystems* and environmental resources.

See also *Disaster*, *Exposure*, *Risk*, and *Vulnerability*.

Heatwave

A period of abnormally hot weather. Heatwaves and warm spells have various and in some cases overlapping definitions.

See also *Extreme weather event*.

Holocene

The current interglacial geological epoch, the second of two epochs within the Quaternary period, the preceding being the Pleistocene. The International Commission on Stratigraphy defines the start of the Holocene at 11,700 years before 2000 (ICS, 2019).

See also *Anthropocene*.

Human behaviour

The way in which a person acts in response to a particular situation or stimulus. Human actions are relevant at different levels, from international, national, and *sub-national actors*, to NGO, firm-level actors, and communities, households, and individual actions.

Adaptation behaviour

Human actions that directly or indirectly affect the risks of climate change *impacts*.

Mitigation behaviour

Human actions that directly or indirectly influence *mitigation*.

Human behavioural change

A transformation or modification of human actions. Behaviour change efforts can be planned in ways that mitigate *climate change* and/or reduce negative consequences of climate change *impacts*.

Human rights

Rights that are inherent to all human beings, universal, inalienable, and indivisible, typically expressed and guaranteed by law. They include the right to life, economic, social, and cultural rights, and the right to development and self-determination (based upon the definition by the UN Office of the High Commissioner).

Procedural rights

Rights to a legal procedure to enforce substantive rights.

Substantive rights

Basic human rights, including the right to the substance of being human such as life itself, liberty and happiness.

Human security

A condition that is met when the vital core of human lives is protected, and when people have the freedom and capacity to live with dignity. In the context of *climate change*, the vital core of human lives includes the universal and culturally specific, material and non-material elements necessary for people to act on behalf of their interests and to live with dignity.

Human system

Any system in which human organizations and *institutions* play a major role. Often, but not always, the term is synonymous with society or social system. Systems such as agricultural systems, urban systems, political systems, technological systems, and economic systems are all human systems in the sense applied in this report.

Hydrological cycle

The cycle in which water evaporates from the oceans and the land surface, is carried over the Earth in atmospheric circulation as water vapour, condenses to form clouds, precipitates as rain or snow, which on land can be intercepted by trees and vegetation, potentially accumulating as snow or ice, provides *runoff* on the land surface, infiltrates into soils, recharges groundwater, discharges into streams, and ultimately, flows out into the oceans as rivers, polar glaciers and *ice sheets*, from which it will eventually evaporate again. The various systems involved in the hydrological cycle are usually referred to as hydrological systems.

Ice sheet

An ice body originating on land that covers an area of continental size, generally defined as covering >50,000 km². An ice sheet flows outward from a high central ice plateau with a small average surface slope. The margins usually slope more steeply, and most ice is discharged through fast flowing ice streams or outlet *glaciers*, often into the sea or into ice shelves floating on the sea. There are only two ice sheets in the modern world, one on Greenland and one on Antarctica. The latter is divided into the East Antarctic Ice Sheet (EAIS), the West Antarctic Ice Sheet (WAIS) and the Antarctic Peninsula ice sheet. During glacial periods there were other ice sheets.

See also *Glacier*.

Impacts (consequences, outcomes)

The consequences of realised risks on natural and *human systems*, where risks result from the interactions of climate-related *hazards* (including *extreme weather and climate events*), *exposure*, and *vulnerability*. Impacts generally refer to effects on lives, *livelihoods*, health and *wellbeing*, *ecosystems* and species, economic, social and cultural assets, services (including *ecosystem services*), and infrastructure. Impacts may be referred to as consequences or outcomes, and can be adverse or beneficial.

See also *Adaptation, Exposure, Hazard, Loss and Damage, and loss and damages*, and *Vulnerability*.

(climate change) Impact assessment

The practice of identifying and evaluating, in monetary and/or non-monetary terms, the effects of climate change on natural and human systems.

Incremental adaptation

See *Adaptation*.

Indigenous knowledge

The understandings, skills and philosophies developed by societies with long histories of interaction with their natural surroundings. For many Indigenous peoples, Indigenous knowledge informs decision-making about fundamental aspects of life, from day-to-day activities to longer term actions. This knowledge is integral to cultural complexes, which also encompass language, systems of classification, resource use practices, social interactions, values, ritual and spirituality. These distinctive ways of knowing are important facets of the world's cultural diversity. (UNESCO, 2018)

See also *Local knowledge*.

Indirect land-use change

See *Land-use change*.

Industrial revolution

A period of rapid industrial growth with far-reaching social and economic consequences, beginning in Britain during the second half of the 18th century and spreading to Europe and later to other countries including the United States. The invention of the steam engine was an important trigger of this development. The industrial revolution marks the beginning of a strong increase in the use of *fossil fuels*, initially coal, and hence emission of *carbon dioxide (CO₂)*.

See also *Pre-industrial*.

Industrialised/developed/developing countries

There are a diversity of approaches for categorising countries on the basis of their level of economic development, and for defining terms such as industrialised, developed, or developing. Several categorisations are used in this report. (1) In the United Nations system, there is no established convention for designating of developed and developing countries or areas. (2) The United Nations Statistics Division specifies developed and developing regions based on common practice. In addition, specific countries are designated as Least Developed Countries (LDC), landlocked developing countries, *small island developing states*, and transition economies. Many countries appear in more than one of these categories. (3) The World Bank uses income as the main criterion for classifying countries as low, lower middle, upper middle, and high income. (4) The UNDP aggregates indicators for life expectancy, educational attainment, and income into a single composite Human Development Index (HDI) to classify countries as low, medium, high, or very high human development.

Inequality

See *Equality*.

Institution

Rules, norms and conventions held in common by social actors that guide, constrain and shape human interaction. Institutions can be formal, such as laws and policies, or informal, such as norms and conventions. Organizations - such as parliaments, regulatory agencies, private firms, and community bodies - develop and act in response to institutional frameworks and the incentives they frame. Institutions can guide, constrain and shape human interaction through direct control, through incentives, and through processes of socialisation.

See also *Institutional capacity*.

Institutional capacity

Building and strengthening individual organisations and providing technical and management training to support integrated planning and decision-making processes between organisations and people, as well as empowerment, social capital, and an enabling environment, including the culture, values and power relations (Willems and Baumert, 2003).

Integrated assessment

A method of analysis that combines results and models from the physical, biological, economic and social sciences and the interactions among these components in a consistent framework to evaluate the status and the consequences of environmental change and the policy responses to it.

See also *Integrated assessment model (IAM)*.

Integrated assessment model (IAM)

Models that integrate knowledge from two or more domains into a single framework. They are one of the main tools for undertaking *integrated assessments*.

One class of IAM used in respect of climate change *mitigation* may include representations of: multiple sectors of the economy, such as energy, *land use* and *land use change*; interactions between sectors; the economy as a whole; associated *greenhouse gas (GHG)* emissions and *sinks*; and reduced representations of the *climate system*. This class of model is used to assess linkages between economic, social and technological development and the evolution of the climate system.

Another class of IAM additionally includes representations of the costs associated with climate change *impacts*, but includes less detailed representations of economic systems. These can be used to assess impacts and mitigation in a *cost-benefit* framework and have been used to estimate the *social cost of carbon*.

Integrated response options

In this report, integrated response options are those options that simultaneously address more than one *land challenge*. These can be categorised into options that rely on a) *land management*, b) value chain management, and c) *risk management*. Integrated response options are not mutually exclusive.

See also *Land challenge*.

Integrated water resources management (IWRM)

A process which promotes the coordinated development and management of water, land and related resources in order to maximise economic and social welfare in an equitable manner without compromising the sustainability of vital *ecosystems*.

Inter-generational equity

See *Equity*.

Internal variability

See *Climate variability*.

Irreversibility

A perturbed state of a dynamical system is defined as irreversible on a given timescale if the recovery timescale from this state due to natural processes is substantially longer than the time it takes for the system to reach this perturbed state.

See also *Tipping point*.

Kyoto Protocol

The Kyoto Protocol to the *United Nations Framework Convention on Climate Change (UNFCCC)* is an international treaty adopted in December 1997 in Kyoto, Japan, at the Third Session of the *Conference of the Parties (COP3)* to the UNFCCC. It contains legally binding commitments, in addition to those included in the UNFCCC. Countries included in Annex B of the Protocol (mostly OECD countries and countries with economies in transition) agreed to reduce their anthropogenic *greenhouse gas (GHG)* emissions (*carbon dioxide (CO₂)*, *methane (CH₄)*, *nitrous oxide (N₂O)*, hydrofluorocarbons (HFCs), perfluorocarbons (PFCs), and sulphur hexafluoride (SF₆)) by at least 5% below 1990 levels in the first commitment period (2008-2012). The Kyoto Protocol entered into force on 16 February 2005 and as of May 2018 had 192 Parties (191 States and the European Union). A second commitment period was agreed in December 2012 at COP18, known as the Doha Amendment to the Kyoto Protocol, in which a new set of Parties committed to reduce GHG emissions by at least 18% below 1990 levels in the period from 2013 to 2020. However, as of May 2018, the Doha Amendment had not received sufficient ratifications to enter into force.

See also *Paris Agreement*.

Land

The terrestrial portion of the biosphere that comprises the natural resources (soil, near surface air, vegetation and other biota, and water), the ecological processes, topography, and human settlements and infrastructure that operate within that system (FAO, 2007; UNCCD, 1994).

Land challenges

In this report, land challenges refers to land-based *mitigation* and *adaptation*, *desertification*, *land degradation* and *food security*.

Land cover

The biophysical coverage of *land* (e.g., bare soil, rocks, forests, buildings and roads or lakes). Land cover is often categorised in broad land-cover classes (e.g., deciduous forest, coniferous forest, mixed forest, grassland, bare ground).

[Note: In some literature assessed in this report, land cover and *land use* are used interchangeably, but the two represent distinct classification systems. For example, the land cover class woodland can be under various land uses such as livestock grazing, recreation, conservation, or wood harvest.]

See also *Land-cover change*, and *Land-use change*.

Land cover change

Change from one *land cover* class to another, due to change in *land use* or change in natural conditions (Pongratz et al., 2018).

See also *Land-use change*, and *Land-management change*.

Land degradation

A negative trend in land condition, caused by direct or indirect human-induced processes including *anthropogenic* climate change, expressed as long-term reduction or loss of at least one of the following: biological productivity, ecological integrity or value to humans.

[Note: This definition applies to *forest* and non-forest land. Changes in land condition resulting solely from natural processes (such as volcanic eruptions) are not considered to be land degradation. Reduction of biological productivity or ecological integrity or value to humans *can* constitute degradation, but any one of these changes need not necessarily be considered degradation.]

Land degradation neutrality

A state whereby the amount and quality of land resources necessary to support *ecosystem* functions and services and enhance *food security* remain stable or increase within specified temporal and spatial scales and ecosystems (UNCCD, 2019).

Land management

Sum of *land-use* practices (e.g., sowing, fertilizing, weeding, harvesting, thinning, clear-cutting) that take place within broader land-use categories. (Pongratz et al., 2018)

Land management change

A change in land management that occurs within a *land-use* category.

See also *Land-use change*.

Land potential

The inherent, long-term potential of the *land* to sustainably generate *ecosystem services*, which reflects the capacity and *resilience* of the land-based natural capital, in the face of ongoing environmental change (UNEP, 2016).

Land rehabilitation

Direct or indirect actions undertaken with the aim of reinstating a level of *ecosystem* functionality, where the goal is provision of goods and services rather than ecological restoration (McDonald, et al., 2016).

Land restoration

The process of assisting the recovery of *land* from a degraded state (McDonald et al., 2016; IPBES, 2018).

Land surface air temperature (LSAT)

The near-surface air temperature over land, typically measured at 1.25-2 m above the ground using standard meteorological equipment.

Land use

The total of arrangements, activities and inputs applied to a parcel of *land*. The term land use is also used in the sense of the social and economic purposes for which land is managed (e.g., grazing, timber extraction, conservation and city dwelling). In national GHG inventories, land use is classified according to the IPCC land use categories of *forest* land, cropland, grassland, *wetlands*, settlements, other lands (see the 2006 IPCC Guidelines for National GHG Inventories for details).

See also *Land-use change* and *Land management*.

Land-use change (LUC)

The change from one *land use* category to another.

[Note: In some of the scientific literature assessed in this report, land-use change encompasses changes in land-use categories as well as changes in *land management*.]

Indirect land-use change (iLUC)

Land use change outside the area of focus, that occurs as a consequence of change in use or management of land within the area of focus, such as through market or policy drivers. For example, if agricultural land is diverted to *biofuel* production, *forest* clearance may occur elsewhere to replace the former agricultural production.

See also *Afforestation, Agriculture, Forestry and Other Land Use (AFOLU), Deforestation, Land use, land-use change and forestry (LULUCF), Reforestation, the IPCC Special Report on Land Use, Land-Use Change, and Forestry (IPCC, 2000,)*, and the *2006 IPCC Guidelines for National GHG Inventories (IPCC, 2006)*.

Land use, land-use change and forestry (LULUCF)

In the context of national *greenhouse gas (GHG)* inventories under the *United Nations Framework Convention on Climate Change (UNFCCC, 2019)*, LULUCF is a GHG inventory sector that covers *anthropogenic* emissions and removals of GHG in managed lands, excluding non-CO₂ agricultural emissions. Following the 2006 IPCC Guidelines for National GHG Inventories, ‘anthropogenic’ land-related GHG fluxes are defined as all those occurring on ‘managed land’, i.e., ‘where human interventions and practices have been applied to perform production, ecological or social functions’. Since managed land may include *carbon dioxide (CO₂)* removals not considered as ‘anthropogenic’ in some of the scientific literature assessed in this report (e.g., removals associated with CO₂ fertilisation and N deposition), the land-related net GHG emission estimates from global models included in this report are not necessarily directly comparable with LULUCF estimates in National GHG Inventories.

See also *Land use change (LUC)*.

Latent heat flux

The turbulent flux of heat from the Earth's surface to the atmosphere that is associated with *evaporation* or condensation of water vapour at the surface; a component of the surface energy budget.

See also *Atmosphere*, and *Flux*.

Leakage

The effects of policies that result in a displacement of the environmental impact, thereby counteracting the intended effects of the initial policies.

Lifecycle assessment (LCA)

Compilation and evaluation of the inputs, outputs and the potential environmental impacts of a product or service throughout its life cycle (ISO, 2018).

Likelihood

The chance of a specific outcome occurring, where this might be estimated probabilistically. Likelihood is expressed in this report using a standard terminology (Mastrandrea et al., 2010).

See also *Agreement, Evidence, Confidence, and Uncertainty*.

Livelihood

The resources used and the activities undertaken in order to live. Livelihoods are usually determined by the entitlements and assets to which people have access. Such assets can be categorised as human, social, natural, physical, or financial.

Local knowledge

The understandings and skills developed by individuals and populations, specific to the places where they live. Local knowledge informs decision-making about fundamental aspects of life, from day-to-day activities to longer term actions. This knowledge is a key element of the social and cultural systems which influence observations of, and responses to *climate change*; it also informs *governance* decisions (UNESCO, 2018).

See also *Indigenous knowledge*.

Lock-in

A situation in which the future development of a system, including infrastructure, technologies, investments, *institutions*, and behavioural norms, is determined or constrained ('locked in') by historic developments.

Long-lived climate forcers (LLCF)

A set of well-mixed *greenhouse gases* with long atmospheric lifetimes. This set of compounds includes *carbon dioxide (CO₂)* and *nitrous oxide (N₂O)*, together with some fluorinated gases. They have a warming effect on *climate*. These compounds accumulate in the atmosphere at decadal to centennial timescales, and their effect on climate hence persists for decades to centuries after their emission. On timescales of decades to a century already emitted emissions of long-lived climate forcers can only be abated by *greenhouse gas removal (GGR)*.

See also *Short-lived climate forcers (SLCF)*.

Loss and Damage, and losses and damages

Research has taken Loss and Damage (capitalised letters) to refer to political debate under the *United Nations Framework Convention on Climate Change (UNFCCC)* following the establishment of the Warsaw Mechanism on Loss and Damage in 2013, which is to 'address loss and damage associated with impacts of climate change, including extreme events and slow onset events, in developing

countries that are particularly vulnerable to the adverse effects of climate change.’ Lowercase letters (losses and damages) have been taken to refer broadly to harm from (observed) *impacts* and (projected) *risks* (Mechler et al., 2018).

Maladaptive actions (Maladaptation)

Actions that may lead to increased *risk* of adverse climate-related outcomes, including via increased *greenhouse gas (GHG)* emissions, increased *vulnerability to climate change*, or diminished welfare, now or in the future. Maladaptation is usually an unintended consequence.

Malnutrition

Deficiencies, excesses, or imbalances in a person’s intake of energy and/or nutrients. The term malnutrition addresses three broad groups of conditions: undernutrition, which includes wasting (low weight-for-height), stunting (low height-for-age) and underweight (low weight-for-age); micronutrient-related malnutrition, which includes micronutrient deficiencies (a lack of important vitamins and minerals) or micronutrient excess; and overweight, obesity and diet-related noncommunicable diseases (such as heart disease, stroke, diabetes and some cancers) (WHO, 2018).

Micronutrient deficiencies are sometimes termed ‘hidden hunger’ to emphasise that people can be malnourished in the sense of deficient without being deficient in calories. Hidden hunger can apply even where people are obese.

Managed forest

Forests subject to human interventions (notably silvicultural management such as planting, pruning, thinning), timber and fuelwood harvest, protection (fire suppression, insect suppression) and management for amenity values or conservation, with defined geographical boundaries (Ogle et al., 2018).

[Note: For a discussion of the term ‘forest’ in the context of National GHG inventories, see the 2006 IPCC Guidelines for National GHG Inventories.]

See also *Managed land*.

Managed grassland

Grasslands on which human interventions are carried out, such as grazing domestic livestock or hay removal.

Managed land

In the context of national *greenhouse gas (GHG)* inventories under the *United Nations Framework Convention on Climate Change (UNFCCC)*, the 2006 IPCC Guidelines for National GHG Inventories (IPCC, 2006) defines managed land ‘where human interventions and practices have been applied to perform production, ecological or social functions’. The IPCC (2006) defines *anthropogenic* GHG emissions and removals in the *LULUCF* sector as all those occurring on ‘managed land’. The key rationale for this approach is that the preponderance of anthropogenic effects occurs on managed lands.

[Note: More details can be found in IPCC 2006 Guidelines for National GHG Inventories, Volume 4, Chapter 1.]

Market failure

When private decisions are based on market prices that do not reflect the real scarcity of goods and services but rather reflect market distortions, they do not generate an efficient allocation of resources but cause welfare losses. A market distortion is any event in which a market reaches a market clearing price that is substantially different from the price that a market would achieve while operating under conditions of perfect competition and state enforcement of legal contracts and the ownership of private property. Examples of factors causing market prices to deviate from real economic scarcity are environmental externalities, public goods, monopoly power, information asymmetry, transaction costs, and non-rational behaviour.

Measurement, reporting and verification (MRV)

Measurement

‘The process of data collection over time, providing basic datasets, including associated accuracy and precision, for the range of relevant variables. Possible data sources are field measurements, field observations, detection through remote sensing and interviews’ (UN REDD, 2009).

Reporting

‘The process of formal reporting of assessment results to the UNFCCC, according to predetermined formats and according to established standards, especially the Intergovernmental Panel on Climate Change (IPCC) Guidelines and GPG (Good Practice Guidance)’ (UN REDD, 2009).

Verification

‘The process of formal verification of reports, for example, the established approach to verify national communications and national inventory reports to the UNFCCC’ (UN REDD, 2009).

Megadrought

See *Drought*.

Methane (CH₄)

One of the six *greenhouse gases (GHGs)* to be mitigated under the *Kyoto Protocol*. Methane is the major component of natural gas and associated with all hydrocarbon fuels. Significant *anthropogenic* emissions also occur as a result of animal husbandry and paddy rice production. Methane is also produced naturally where organic matter decays under anaerobic conditions, such as in *wetlands*.

Migrant

See *Migration*.

Migration

‘The movement of a person or a group of persons, either across an international border, or within a State. It is a population movement, encompassing any kind of movement of people, whatever its length, composition and causes; it includes migration of refugees, displaced persons, economic migrants, and persons moving for other purposes, including family reunification’ (IOM, 2018).

Migrant

‘Any person who is moving or has moved across an international border or within a State away from his/her habitual place of residence, regardless of (1) the person’s legal status; (2) whether the movement is voluntary or involuntary; (3) what the causes for the movement are; or (4) what the length of the stay is’ (IOM, 2018).

See also *(Internal) Displacement*.

Millennium Development Goals (MDGs)

A set of eight time-bound and measurable goals for combating *poverty*, hunger, disease, illiteracy, discrimination against women and environmental degradation. These goals were agreed at the UN Millennium Summit in 2000 together with an action plan to reach the goals by 2015.

Mitigation (of climate change)

A human intervention to reduce emissions or enhance the *sinks* of *greenhouse gases*.

Mitigation behaviour

See *Human behaviour*.

Mitigation measures

In climate policy, *mitigation* measures are technologies, processes or practices that contribute to mitigation, for example renewable energy technologies, waste minimization processes, public transport commuting practices.

Mitigation option

A technology or practice that reduces *greenhouse gas (GHG)* emissions or enhances *sinks*.

Mitigation pathways

See *Pathways*.

Mitigation scenario

A plausible description of the future that describes how the (studied) system responds to the implementation of *mitigation* policies and measures.

See also *Emission scenario*, *Pathways*, *Socio-economic scenarios*, and *Stabilisation (of GHG or CO₂-equivalent concentration)*.

Monitoring and evaluation (M&E)

Mechanisms put in place at national to local scales to respectively monitor and evaluate efforts to reduce *greenhouse gas* emissions and/or adapt to the *impacts* of *climate change* with the aim of systematically identifying, characterizing and assessing progress over time.

Motivation (of an individual)

An individual's reason or reasons for acting in a particular way; individuals may consider various consequences of actions, including financial, social, affective, and environmental consequences. Motivation can arise from factors external or internal to the individual.

Multi-level governance

See *Governance*.

Narratives (in the context of scenarios)

Qualitative descriptions of plausible future world evolutions, describing the characteristics, general logic and developments underlying a particular quantitative set of *scenarios*. Narratives are also referred to in the literature as ‘storylines’.

See also *Scenario*, *Scenario storyline*, and *Pathways*.

Nationally Determined Contributions (NDCs)

A term used under the *United Nations Framework Convention on Climate Change (UNFCCC)* whereby a country that has joined the *Paris Agreement* outlines its plans for reducing its emissions. Some countries NDCs also address how they will adapt to climate change impacts, and what support they need from, or will provide to, other countries to adopt low-carbon pathways and to build climate resilience. According to Article 4 paragraph 2 of the *Paris Agreement*, each Party shall prepare, communicate and maintain successive NDCs that it intends to achieve. In the lead up to 21st *Conference of the Parties* in Paris in 2015, countries submitted Intended Nationally Determined Contributions (INDCs). As countries join the Paris Agreement, unless they decide otherwise, this INDC becomes their first Nationally Determined Contribution (NDC).

Negative emissions

Removal of *greenhouse gases (GHGs)* from the *atmosphere* by deliberate human activities, i.e., in addition to the removal that would occur via natural *carbon cycle* processes.

See also *Net negative emissions*, *Net-zero emissions*, *Carbon dioxide removal (CDR)*, and *Greenhouse gas removal (GGR)*.

Negative emissions technologies

An activity or mechanism that results in *negative emissions*.

Net negative emissions

A situation of net negative emissions is achieved when, as result of human activities, more *greenhouse gases (GHG)* are removed from the *atmosphere* than are emitted into it. Where multiple greenhouse gases are involved, the quantification of *negative emissions* depends on the climate metric chosen to compare emissions of different gases (such as global warming potential, global temperature change potential, and others, as well as the chosen time horizon).

See also *Negative emissions*, *Net-zero emissions* and *Net-zero CO₂ emissions*.

Net-zero CO₂ emissions

Conditions in which any remaining *anthropogenic* carbon dioxide (CO₂) emissions are balanced by anthropogenic CO₂ removals over a specified period.

See also *Net-zero emissions*, and *Net negative emissions*.

Net-zero emissions

Net-zero emissions are achieved when emissions of *greenhouse gases (GHGs)* to the *atmosphere* are balanced by *anthropogenic removals*. Where multiple greenhouse gases are involved, the quantification of net-zero emissions depends on the climate metric chosen to compare emissions of different gases (such as global warming potential, global temperature change potential, and others, as well as the chosen time horizon).

See also *Net-zero CO₂ emissions*, *Negative emissions*, and *Net negative emission*.

Nitrous oxide (N₂O)

One of the six *greenhouse gases (GHGs)* to be mitigated under the *Kyoto Protocol*. The main *anthropogenic* source of N₂O is agriculture (soil and animal manure management), but important contributions also come from sewage treatment, *fossil fuel* combustion, and chemical industrial processes. N₂O is also produced naturally from a wide variety of biological sources in soil and water, particularly microbial action in wet tropical *forests*.

Non-overshoot pathways

See *Pathways*.

Nutrition transition

A predictable change in *dietary patterns* associated with a country's economic development whereby 'problems of under- and overnutrition often coexist, reflecting the trends in which an increasing proportion of people consume the types of diets associated with a number of chronic diseases' (Popkin, 1994).

Ocean acidification (OA)

A reduction in the *pH* of the ocean, accompanied by other chemical changes, over an extended period, typically decades or longer, which is caused primarily by uptake of *carbon dioxide (CO₂)* from the *atmosphere*, but can also be caused by other chemical additions or subtractions from the ocean. *Anthropogenic* ocean acidification refers to the component of pH reduction that is caused by human activity (IPCC, 2011, p. 37).

See also *Climate change*.

Ocean fertilisation

Deliberate increase of nutrient supply to the near-surface ocean in order to enhance biological production through which additional *carbon dioxide (CO₂)* from the *atmosphere* is sequestered. This can be achieved by the addition of micro-nutrients or macro-nutrients. Ocean fertilisation is regulated by the London Protocol.

Overshoot

See *Temperature overshoot*.

Overshoot pathways

See *Pathways*.

Ozone (O₃)

The triatomic form of oxygen (O₃). In the *troposphere*, it is created both naturally and by photochemical reactions involving gases resulting from human activities (smog). Tropospheric ozone acts as a *greenhouse gas (GHG)*. In the *stratosphere*, it is created by the interaction between solar ultraviolet radiation and molecular oxygen (O₂). Stratospheric ozone plays a dominant role in the stratospheric radiative balance. Its concentration is highest in the ozone layer.

Paris Agreement

The Paris Agreement under the *United Nations Framework Convention on Climate Change (UNFCCC)* was adopted on December 2015 in Paris, France, at the 21st session of the *Conference of the Parties (COP)* to the UNFCCC. The agreement, adopted by 196 Parties to the UNFCCC, entered into force on 4 November 2016 and as of May 2018 had 195 Signatories and was ratified by 177 Parties. One of the goals of the Paris Agreement is ‘Holding the increase in the global average temperature to well below 2°C above pre-industrial levels and pursuing efforts to limit the temperature increase to 1.5°C above pre-industrial levels’, recognising that this would significantly reduce the risks and impacts of climate change. Additionally, the Agreement aims to strengthen the ability of countries to deal with the impacts of climate change. The Paris Agreement is intended to become fully effective in 2020.

See also *Kyoto Protocol*, and *Nationally Determined Contributions (NDCs)*.

Participatory governance

See *Governance*.

Pasture

Area covered with grass or other plants used or suitable for grazing of livestock; grassland.

Pathways

The temporal evolution of natural and/or *human systems* towards a future state. Pathway concepts range from sets of quantitative and qualitative *scenarios* or *narratives* of potential futures to solution-oriented decision-making processes to achieve desirable societal goals. Pathway approaches typically focus on biophysical, techno-economic, and/or socio-behavioural trajectories and involve various dynamics, goals, and actors across different scales.

1.5°C pathway

A pathway of emissions of *greenhouse gases* and other climate forcers that provides an approximately one-in-two to two-in-three chance, given current knowledge of the climate response, of *global warming* either remaining below 1.5°C or returning to 1.5°C by around 2100 following an overshoot. See also *Temperature overshoot*.

Adaptation pathways

A series of *adaptation* choices involving trade-offs between short-term and long-term goals and values. These are processes of deliberation to identify solutions that are meaningful to people in the context of their daily lives and to avoid potential maladaptation.

Development pathways

Development pathways are trajectories based on an array of social, economic, cultural, technological, *institutional*, and biophysical features that characterise the interactions between human and natural systems and outline visions for the future, at a particular scale.

Mitigation pathways

A mitigation pathway is a temporal evolution of a set of *mitigation scenario* features, such as *greenhouse gas (GHG)* emissions and socio-economic development.

Overshoot pathways

Pathways that exceed the stabilization level (concentration, *forcing*, or temperature) before the end of a time horizon of interest (e.g., before 2100) and then decline towards that level by that time. Once the target level is exceeded, removal by *sinks* of *greenhouse gases (GHGs)* is required.

See also *Temperature overshoot*.

Non-overshoot pathways

Pathways that stay below the stabilization level (concentration, *forcing*, or temperature) during the time horizon of interest (e.g., until 2100).

Representative Concentration Pathways (RCPs)

Scenarios that include time series of emissions and concentrations of the full suite of *greenhouse gases (GHGs)* and *aerosols* and chemically active gases, as well as *land use/land cover* (Moss et al., 2008). The word representative signifies that each RCP provides only one of many possible scenarios that would lead to the specific *radiative forcing* characteristics. The term pathway emphasises the fact that not only the long-term concentration levels, but also the trajectory taken over time to reach that outcome are of interest (Moss et al., 2010). RCPs were used to develop *climate projections* in CMIP5.

- RCP2.6: One pathway where radiative forcing peaks at approximately 3 W m^{-2} and then declines to be limited at 2.6 W m^{-2} in 2100 (the corresponding Extended Concentration Pathway, or ECP, has constant emissions after 2100).
- RCP4.5 and RCP6.0: Two intermediate stabilisation pathways in which radiative forcing is limited at approximately 4.5 W m^{-2} and 6.0 W m^{-2} in 2100 (the corresponding ECPs have constant concentrations after 2150).
- RCP8.5: One high pathway which leads to $>8.5 \text{ W m}^{-2}$ in 2100 (the corresponding ECP has constant emissions after 2100 until 2150 and constant concentrations after 2250).

See also *Coupled Model Intercomparison Project (CMIP)*, and *Shared Socio-economic Pathways (SSPs)*.

Shared Socio-economic Pathways (SSPs)

Shared Socio-economic Pathways (SSPs) were developed to complement the Representative Concentration Pathways (RCPs) with varying socio-economic challenges to *adaptation* and *mitigation* (O'Neill et al., 2014). Based on five *narratives*, the SSPs describe alternative socio-economic futures in the absence of climate policy intervention, comprising sustainable development (SSP1), regional rivalry (SSP3), inequality (SSP4), fossil-fueled development (SSP5), and a middle-of-the-road development (SSP2) (O'Neill et al., 2017; Riahi et al., 2017). The combination of SSP-based socio-economic scenarios and RCP-based *climate projections* provides an integrative frame for climate *impact* and policy analysis.

Transformation pathways

Trajectories describing consistent sets of possible futures of *greenhouse gas (GHG)* emissions, atmospheric concentrations, or *global mean surface temperatures* implied from *mitigation* and *adaptation* actions associated with a set of broad and irreversible economic, technological, societal, and behavioural changes. This can encompass changes in the way energy and infrastructure are used and produced, natural resources are managed and *institutions* are set up and in the pace and direction of technological change (TC).

See also *Scenario*, *Scenario storyline*, *Emission scenario*, *Mitigation scenario*, *Baseline scenario*, *Stabilisation (of GHG or CO₂-equivalent concentration)*, and *Narratives*.

Peat

Soft, porous or compressed, sedentary deposit of which a substantial portion is partly decomposed plant material with high water content in the natural state (up to about 90 percent) (IPCC, 2013).

See also *Peatlands*.

Peatlands

Peatland is a land where soils are dominated by *peat*.

See also *Reservoir*, and *Sink*.

Peri-urban areas

Parts of a city that appear to be quite rural but are in reality strongly linked functionally to the city in its daily activities.

Permafrost

Ground (soil or rock and included ice and organic material) that remains at or below 0°C for at least two consecutive years.

pH

A dimensionless measure of the acidity of a solution given by its concentration of hydrogen ions ($[H^+]$). pH is measured on a logarithmic scale where $pH = -\log_{10}[H^+]$. Thus, a pH decrease of 1 unit corresponds to a 10-fold increase in the concentration of H^+ , or acidity.

See also *Ocean acidification*.

Phenology

The relationship between biological phenomena that recur periodically (e.g., development stages, migration) and *climate* and seasonal changes.

Planetary health

The Rockefeller-Lancet Commission defines planetary health as ‘the achievement of the highest attainable standard of health, wellbeing, and equity worldwide through judicious attention to the human systems — political, economic, and social — that shape the future of humanity *and* the Earth’s natural systems that define the safe environmental limits within which humanity can flourish. Put simply, planetary health is the health of human civilisation and the state of the natural systems on which it depends’ (Whitmee et al., 2015).

Political economy

The set of interlinked relationships between people, the state, society and markets as defined by law, politics, economics, customs and power that determine the outcome of trade and transactions and the distribution of wealth in a country or economy.

Poverty

A complex concept with several definitions stemming from different schools of thought. It can refer to material circumstances (such as need, pattern of deprivation or limited resources), economic conditions (such as standard of living, *inequality* or economic position) and/or social relationships (such as social class, dependency, exclusion, lack of basic security or lack of entitlement).

See also *Poverty eradication*.

Poverty eradication

A set of measures to end *poverty* in all its forms everywhere.

See also *Sustainable Development Goals (SDGs)*.

Pool, carbon and nitrogen

A *reservoir* in the earth system where elements, such as carbon and nitrogen, reside in various chemical forms for a period of time.

Precursors

Atmospheric compounds that are not *greenhouse gases* (GHGs) or *aerosols*, but that have an effect on GHG or aerosol concentrations by taking part in physical or chemical processes regulating their production or destruction rates.

See also *Aerosol*, and *Greenhouse gas (GHG)*.

Pre-industrial

The multi-century period prior to the onset of large-scale industrial activity around 1750. The *reference period* 1850-1900 is used to approximate pre-industrial *global mean surface temperature (GMST)*.

See also *Industrial revolution*.

Primary production

The synthesis of organic compounds by plants and microbes, on land or in the ocean, primarily by photosynthesis using light and *carbon dioxide (CO₂)* as sources of energy and carbon respectively. It can also occur through chemosynthesis, using chemical energy, e.g., in deep sea vents.

Gross Primary Production (GPP)

The total amount of carbon fixed by photosynthesis over a specific time period.

Net primary production (NPP)

The amount of carbon fixed by photosynthesis minus the amount lost by respiration over a specified time period.

Procedural equity

See *Equity*.

Procedural rights

See *Human rights*.

Projection

A potential future evolution of a quantity or set of quantities, often computed with the aid of a model. Unlike predictions, projections are conditional on assumptions concerning, for example, future socio-economic and technological developments that may or may not be realised.

See also *Climate projection*, *Scenario*, and *Pathways*.

Radiative forcing

The change in the net, downward minus upward, radiative flux (expressed in W m^{-2}) at the tropopause or top of *atmosphere* due to a change in an driver of *climate change*, such as a change in the concentration of *carbon dioxide (CO₂)*, the concentration of volcanic aerosols or the output of the

Sun. The traditional radiative forcing is computed with all tropospheric properties held fixed at their unperturbed values, and after allowing for stratospheric temperatures, if perturbed, to readjust to radiative-dynamical equilibrium. Radiative forcing is called instantaneous if no change in stratospheric temperature is accounted for. The radiative forcing once rapid adjustments are accounted for is termed the effective radiative forcing. Radiative forcing is not to be confused with cloud radiative forcing, which describes an unrelated measure of the impact of clouds on the radiative flux at the top of the atmosphere.

Reasons for concern (RFCs)

Elements of a classification framework, first developed in the IPCC Third Assessment Report, which aims to facilitate judgments about what level of *climate change* may be dangerous (in the language of Article 2 of the *UNFCCC*) by aggregating risks from various sectors, considering hazards, exposures, vulnerabilities, capacities to adapt, and the resulting impacts.

Reducing Emissions from Deforestation and Forest Degradation (REDD+)

REDD+ refers to reducing emissions from *deforestation*; reducing emissions from forest degradation; conservation of forest carbon stocks; sustainable management of forests; and enhancement of forest carbon stocks (see UNFCCC decision 1/CP.16, para. 70).

Reference period

The period relative to which *anomalies* are computed.

See also *Anomalies*.

Reference scenario

See *Baseline scenario*.

Reforestation

Conversion to *forest of land* that has previously contained forests but that has been converted to some other use.

[Note: For a discussion of the term forest and related terms such as *afforestation*, *reforestation* and *deforestation* in the context of reporting and accounting Article 3.3 and 3.4 activities under the *Kyoto Protocol*, see 2013 Revised Supplementary Methods and Good Practice Guidance Arising from the Kyoto Protocol.]

See also *Afforestation*, *Deforestation*, and *Reducing Emissions from Deforestation and Forest Degradation (REDD+)*.

Region

A relatively large-scale land or ocean area characterised by specific geographical and climatological features. The *climate* of a land-based region is affected by regional and local scale features like topography, *land use* characteristics and large water bodies, as well as remote influences from other regions, in addition to global climate conditions. The IPCC defines a set of standard regions for analyses of observed climate trends and climate model projections (see Fig. 3.2; AR5, SREX).

Remaining carbon budget

Cumulative global *carbon dioxide* (CO_2) emissions from the start of 2018 to the time that CO_2 emissions reach net-zero that would result, at some probability, in limiting *global warming* to a given level, accounting for the impact of other *anthropogenic emissions*.

See also *Carbon budget*.

Representative concentration pathways (RCPs)

See *Pathways*

Reservoir

A component or components of the climate system where a *greenhouse gas* (*GHG*) or a *precursor* of a greenhouse gas is stored (UNFCCC Article 1.7).

Resilience

The capacity of interconnected social, economic and ecological systems to cope with a hazardous event, trend or disturbance, responding or reorganizing in ways that maintain their essential function, identity and structure. Resilience is a positive attribute when it maintains capacity for *adaptation*, learning and/or *transformation* (adapted from the Arctic Council, 2013).

See also *Hazard*, *Risk*, and *Vulnerability*.

Respiration

The process whereby living organisms convert organic matter to *carbon dioxide* (CO_2), releasing energy and consuming molecular oxygen.

Rewetting

‘The deliberate action of changing a drained soil into a wet soil, e.g. by blocking *drainage* ditches, disabling pumping facilities or breaching obstructions’ (IPCC, 2013).

See also *Drainage*.

Risk

The potential for adverse consequences for human or ecological systems, recognizing the diversity of values and objectives associated with such systems. In the context of *climate change*, risks can arise from potential *impacts* of climate change as well as human responses to climate change. Relevant adverse consequences include those on lives, *livelihoods*, health and *wellbeing*, economic, social and cultural assets and investments, infrastructure, services (including *ecosystem services*), *ecosystems* and species.

In the context of climate change impacts, risks result from dynamic interactions between climate-related *hazards* with the *exposure* and *vulnerability* of the affected human or ecological system to the hazards. Hazards, exposure and vulnerability may each be subject to *uncertainty* in terms of magnitude and likelihood of occurrence, and each may change over time and space due to socio-economic changes and human decision-making (see also *risk management*, *adaptation*, and *mitigation*).

In the context of climate change responses, risks result from the potential for such responses not achieving the intended objective(s), or from potential trade-offs with, or negative side-effects on, other societal objectives, such as the *Sustainable Development Goals* (see also *risk trade-off*). Risks

can arise for example from uncertainty in implementation, effectiveness or outcomes of climate policy, climate-related investments, technology development or adoption, and system *transitions*.

Risk assessment

The qualitative and/or quantitative scientific estimation of *risks*.

See also *Risk management*, and *Risk perception*.

Risk management

Plans, actions, strategies or policies to reduce the *likelihood* and/or magnitude of adverse potential consequences, based on assessed or perceived *risks*.

See also *Risk assessment*, and *Risk perception*.

Risk perception

The subjective judgment that people make about the characteristics and severity of a *risk*.

See also *Risk assessment*, and *Risk management*.

Risk trade-off

The change in portfolio of *risks* that occurs when a countervailing risk is generated (knowingly or inadvertently) by an intervention to reduce the target risk (Wiener and Graham, 2009).

See also *Adverse side-effect*, and *Co-benefits*.

Runoff

The flow of water over the surface or through the subsurface, which typically originates from the part of liquid precipitation and/or snow/ice melt that does not evaporate or refreeze, and is not transpired.

See also *Hydrological cycle*.

Saline soils

Soils with levels of soluble salts (commonly sulphates and chlorides of calcium and magnesium) in the saturation extract high enough to negatively affect plant growth. Saline soils are usually flocculated and have good water permeability (Well and Brady, 2016).

See also *Soil salinity* and *Sodic soils*.

Scenario

A plausible description of how the future may develop based on a coherent and internally consistent set of assumptions about key driving forces (e.g., rate of technological change (TC), prices) and relationships. Note that scenarios are neither predictions nor forecasts, but are used to provide a view of the implications of developments and actions.

See also *Baseline scenario*, *Emission scenario*, *Mitigation scenario* and *Pathways*.

Scenario storyline

A *narrative* description of a *scenario* (or family of scenarios), highlighting the main scenario characteristics, relationships between key driving forces and the dynamics of their evolution. Also referred to as ‘narratives’ in the scenario literature.

Sea ice

Ice found at the sea surface that has originated from the freezing of seawater. Sea ice may be discontinuous pieces (ice floes) moved on the ocean surface by wind and currents (pack ice), or a motionless sheet attached to the coast (land-fast ice). Sea ice concentration is the fraction of the ocean covered by ice. Sea ice less than one year old is called first-year ice. Perennial ice is sea ice that survives at least one summer. It may be subdivided into second-year ice and multi-year ice, where multiyear ice has survived at least two summers.

Sea level change (sea level rise/sea level fall)

Change to the height of sea level, both globally and locally (relative sea level change) due to (1) a change in ocean volume as a result of a change in the mass of water in the ocean, (2) changes in ocean volume as a result of changes in ocean water density, (3) changes in the shape of the ocean basins and changes in the Earth’s gravitational and rotational fields, and (4) local subsidence or uplift of the land. Global mean sea level change resulting from change in the mass of the ocean is called barystatic. The amount of barystatic sea level change due to the addition or removal of a mass of water is called its sea level equivalent (SLE). Sea level changes, both globally and locally, resulting from changes in water density are called steric. Density changes induced by temperature changes only are called thermosteric, while density changes induced by salinity changes are called halosteric. Barystatic and steric sea level changes do not include the effect of changes in the shape of ocean basins induced by the change in the ocean mass and its distribution.

Sea surface temperature (SST)

The subsurface bulk temperature in the top few meters of the ocean, measured by ships, buoys, and drifters. From ships, measurements of water samples in buckets were mostly switched in the 1940s to samples from engine intake water. Satellite measurements of skin temperature (uppermost layer; a fraction of a millimetre thick) in the infrared or the top centimetre or so in the microwave are also used, but must be adjusted to be compatible with the bulk temperature.

Sendai Framework for Disaster Risk Reduction

The Sendai Framework for Disaster Risk Reduction 2015-2030 outlines seven clear targets and four priorities for action to prevent new, and to reduce existing disaster risks. The voluntary, non-binding agreement recognises that the State has the primary role to reduce disaster risk but that responsibility should be shared with other stakeholders including local government, the private sector and other stakeholders, with the aim for the substantial reduction of disaster risk and losses in lives, *livelihoods* and health and in the economic, physical, social, cultural and environmental assets of persons, businesses, communities and countries.

Sequestration

See *Uptake* and *Carbon sequestration*.

Shared socio-economic pathways (SSPs)

See *Pathways*.

Short-lived climate forcers (SLCF)

A set of compounds that are primarily composed of those with short lifetimes in the atmosphere compared to well-mixed *greenhouse gases (GHGs)*, and are also referred to as near-term climate forcers. This set of compounds includes *methane (CH₄)*, which is also a well-mixed greenhouse gas, as well as *ozone (O₃)* and *aerosols*, or their *precursors*, and some halogenated species that are not well-mixed greenhouse gases. These compounds do not accumulate in the atmosphere at decadal to centennial timescales, and so their effect on *climate* is predominantly in the first decade after their emission, although their changes can still induce long-term climate effects such as *sea level change*. Their effect can be cooling or warming. A subset of exclusively warming short-lived climate forcers is referred to as short-lived climate pollutants.

See also *Long-lived climate forcers (LLCF)*.

Short-lived climate pollutants (SLCP)

See *Short-lived climate forcers (SLCF)*.

Sink

Any process, activity or mechanism which removes a *greenhouse gas*, an *aerosol* or a *precursor* of a *greenhouse gas* from the *atmosphere* (UNFCCC Article 1.8).

See also *Sequestration*, *Source*, and *Uptake*.

Small Island Developing States (SIDS)

Small Island Developing States (SIDS), as recognised by the United Nations OHRLLS (Office of the High Representative for the Least Developed Countries, Landlocked Developing Countries and Small Island Developing States), are a distinct group of developing countries facing specific social, economic and environmental vulnerabilities (UN-OHRLLS, 2011). They were recognised as a special case both for their environment and development at the Rio Earth Summit in Brazil in 1992. Fifty eight countries and territories are presently classified as SIDS by the UN OHRLLS, with 38 being UN member states and 20 being Non-UN Members or Associate Members of the Regional Commissions (UN-OHRLLS, 2018).

Social costs

The full costs of an action in terms of social welfare losses, including external costs associated with the impacts of this action on the environment, the economy (GDP, employment) and on the society as a whole.

Social cost of carbon (SCC)

The net present value of aggregate climate damages (with overall harmful damages expressed as a number with positive sign) from one more tonne of carbon in the form of *carbon dioxide (CO₂)*, conditional on a global emissions trajectory over time.

Social-ecological system

An integrated system that includes human societies and *ecosystems*, in which humans are part of nature. The functions of such a system arise from the interactions and interdependence of the social and ecological subsystems. The system's structure is characterised by reciprocal feedbacks, emphasising that humans must be seen as a part of, not apart from, nature (Artic Council, 2016; Berkes and Folke, 1998).

Social inclusion

A process of improving the terms of participation in society, particularly for people who are disadvantaged, through enhancing opportunities, access to resources, and respect for rights (UN, DESA 2016).

Social learning

A process of social interaction through which people learn new behaviours, capacities, values, and attitudes.

Societal (social) transformation

See *Transformation*.

Socio-economic scenario

A *scenario* that describes a possible future in terms of population, *gross domestic product (GDP)*, and other socio-economic factors relevant to understanding the implications of *climate change*.

See also *Baseline scenario*, *Emission scenario*, *Mitigation scenario*, and *Pathways*.

Socio-technical transitions

Where technological change is associated with social systems and the two are inextricably linked.

Soil carbon sequestration (SCS)

Land management changes which increase the *soil organic carbon* content, resulting in a net removal of *carbon dioxide (CO₂)* from the *atmosphere*.

Soil conservation

The maintenance of soil fertility through controlling erosion, preserving soil organic matter, ensuring favourable soil physical properties, and retaining nutrients (Young, 1989).

Soil erosion

The displacement of the soil by the action of water or wind. Soil erosion is a major process of *land degradation*.

Soil organic carbon

Carbon contained in *soil organic matter*.

Soil organic matter

The organic component of soil, comprising plant and animal residue at various stages of decomposition, and soil organisms.

Soil moisture

Water stored in the soil in liquid or frozen form. Root-zone soil moisture is of most relevance for plant activity.

Soil salinity

The concentration of soluble salts in the water extracted from a saturated soil (saturation extract), comprising chlorides and sulphates of Sodium (Na^+), calcium (Ca^{2+}) and magnesium (Mg^{2+}) as well as carbonate salts (adapted from FAO, 1985).

See also *Saline soils*, and *Sodic soils*.

Sodic soils

Soils with disproportionately high concentration of sodium (Na^+) in relation to calcium (Ca^{2+}) and magnesium (Mg^{2+}) adsorbed at the cation exchange site on the surface of soil particles. Sodic soils are characterised by a poor soil structure and poor aeration (NDSU, 2014).

See also *Soil salinity*, and *Sodic soils*.

Source

Any process or activity which releases a *greenhouse gas*, an *aerosol* or a *precursor* of a *greenhouse gas* into the *atmosphere* (UNFCCC Article 1.9).

See also *Sink*.

Stabilisation (of GHG or CO₂-equivalent concentration)

A state in which the *atmospheric* concentrations of one *greenhouse gas* (*GHG*) (e.g., carbon dioxide) or of a CO₂-equivalent basket of GHGs (or a combination of GHGs and *aerosols*) remains constant over time.

Stranded assets

Assets exposed to devaluations or conversion to ‘liabilities’ because of unanticipated changes in their initially expected revenues due to innovations and/or evolutions of the business context, including changes in public regulations at the domestic and international levels.

Stratosphere

The highly stratified region of the *atmosphere* above the *troposphere* extending from about 10 km (ranging from 9 km at high latitudes to 16 km in the tropics on average) to about 50 km altitude.

See also *Atmosphere*, and *Troposphere*.

Subnational actors

State/provincial, regional, metropolitan and local/municipal governments as well as non-party stakeholders, such as civil society, the private sector, cities and other subnational authorities, local communities and indigenous peoples.

Substantive rights

See *Human rights*.

Supply-side measures

See *Demand and supply-side measures*.

Surface temperature

See *Global mean surface temperature (GMST)*, *Land surface air temperature*, and *Sea surface temperature (SST)*.

Sustainability

A dynamic process that guarantees the persistence of natural and *human systems* in an equitable manner.

Sustainable development (SD)

Development that meets the needs of the present without compromising the ability of future generations to meet their own needs (WCED, 1987) and balances social, economic and environmental concerns.

See also *Sustainable Development Goals (SDGs)*, and *Development pathways* (under *Pathways*).

Sustainable Development Goals (SDGs)

The 17 global goals for development for all countries established by the United Nations through a participatory process and elaborated in the *2030 Agenda for Sustainable Development*, including ending *poverty* and hunger; ensuring health and *wellbeing*, education, *gender equality*, clean water and energy, and decent work; building and ensuring *resilient* and sustainable infrastructure, cities and consumption; reducing *inequalities*; protecting land and water *ecosystems*; promoting peace, justice and partnerships; and taking urgent action on *climate change*.

See also *Sustainable development (SD)*.

Sustainable forest management

The stewardship and use of *forests* and forest lands in a way, and at a rate, that maintains their *biodiversity*, productivity, regeneration capacity, vitality and their potential to fulfil, now and in the future, relevant ecological, economic and social functions, at local, national, and global levels, and that does not cause damage to other *ecosystems* (Forest Europe, 1993).

Sustainable intensification (of agriculture)

Increasing yields from the same area of land while decreasing negative environmental impacts of agricultural production and increasing the provision of environmental services (CGIAR, 2019).

[Note: this definition is based on the concept of meeting demand from a finite land area, but it is scale-dependent. Sustainable intensification at a given scale (e.g., global or national) may require a decrease in production intensity at smaller scales and in particular places (often associated with previous, unsustainable, intensification) to achieve *sustainability* (Garnett et al., 2013).]

Sustainable land management

The stewardship and use of *land* resources, including soils, water, animals and plants, to meet changing human needs, while simultaneously ensuring the long-term productive potential of these resources and the maintenance of their environmental functions (Adapted from WOCAT, undated).

Technology transfer

The exchange of knowledge, hardware and associated software, money and goods among stakeholders, which leads to the spread of technology for *adaptation* or *mitigation*. The term encompasses both diffusion of technologies and technological cooperation across and within countries.

Teleconnections

A statistical association between climate variables at widely separated, geographically-fixed spatial locations. Teleconnections are caused by large spatial structures such as basin-wide coupled modes of ocean-atmosphere variability, Rossby wave-trains, mid-latitude jets and storm tracks, etc.

Temperature overshoot

The temporary exceedance of a specified level of *global warming*, such as 1.5°C. Overshoot implies a peak followed by a decline in global warming, achieved through *anthropogenic removal* of *carbon dioxide (CO₂)* exceeding remaining CO₂ emissions globally.

See also *Pathways (Subterms: Overshoot pathways, Non-overshoot Pathways)*.

Tier

In the context of the IPCC Guidelines for National Greenhouse Gas Inventories, a tier represents a level of methodological complexity. Usually three tiers are provided. Tier 1 is the basic method, Tier 2 intermediate and Tier 3 most demanding in terms of complexity and data requirements. Tiers 2 and 3 are sometimes referred to as higher tier methods and are generally considered to be more accurate (IPCC, 2019).

Tipping point

A level of change in system properties beyond which a system reorganises, often abruptly, and does not return to the initial state even if the drivers of the change are abated. For the *climate system*, it refers to a critical threshold beyond which global or regional climate changes from one stable state to another stable state. Tipping points are also used when referring to *impact*: the term can imply that an impact tipping point is (about to be) reached in a natural or *human system*.

See also *Irreversibility*.

Transformation

A change in the fundamental attributes of natural and human systems.

Societal (social) transformation

A profound and often deliberate shift initiated by communities toward sustainability, facilitated by changes in individual and collective values and behaviours, and a fairer balance of political, cultural, and *institutional* power in society.

Transformation pathways

See *Pathways*.

Transformational adaptation

See *Adaptation*.

Transformative change

A system wide change that alters the fundamental attributes of the system.

Transient climate response to cumulative CO₂ emissions (TCRE)

The transient global average surface temperature change per unit cumulative *carbon dioxide* (CO₂) emissions, usually 1000 GtC. TCRE combines both information on the airborne fraction of cumulative CO₂ emissions (the fraction of the total CO₂ emitted that remains in the *atmosphere*, which is determined by *carbon cycle* processes) and on the *transient climate response* (TCR).

See also *Transient climate response (TCR)* (under *Climate sensitivity*).

Transit-oriented development (TOD)

An approach urban development that maximises the amount of residential, business and leisure space within walking distance of efficient public transport, so as to enhance mobility of citizens, the viability of public transport and the value of urban land in mutually supporting ways.

Transition

The process of changing from one state or condition to another in a given period of time. Transition can occur in individuals, firms, cities, regions and nations, and can be based on incremental or transformative change.

Tropical cyclone

The general term for a strong, cyclonic-scale disturbance that originates over tropical oceans. Distinguished from weaker systems (often named tropical disturbances or depressions) by exceeding a threshold wind speed. A tropical storm is a tropical cyclone with one-minute average surface winds between 18 and 32 m s⁻¹. Beyond 32 m s⁻¹, a tropical cyclone is called a hurricane, typhoon, or cyclone, depending on geographic location.

See also *Extratropical cyclone*.

Troposphere

The lowest part of the *atmosphere*, from the surface to about 10 km in altitude at mid-latitudes (ranging from 9 km at high latitudes to 16 km in the tropics on average), where clouds and weather phenomena occur. In the troposphere, temperatures generally decrease with height.

See also *Atmosphere*, and *Stratosphere*.

Uncertainty

A state of incomplete knowledge that can result from a lack of information or from disagreement about what is known or even knowable. It may have many types of sources, from imprecision in the data to ambiguously defined concepts or terminology, incomplete understanding of critical processes, or uncertain *projections* of *human behaviour*. Uncertainty can therefore be represented by quantitative measures (e.g., a probability density function) or by qualitative statements (e.g., reflecting the judgment of a team of experts) (see IPCC, 2004; Mastrandrea et al., 2010; Moss and Schneider, 2000).

See also *Confidence*, and *Likelihood*.

United Nations Convention to Combat Desertification (UNCCD)

A legally binding international agreement linking environment and development to sustainable land management, established in 1994. The Convention's objective is 'to combat desertification and mitigate the effects of drought in countries experiencing drought and/or desertification'. The Convention specifically addresses the arid, semi-arid and dry sub-humid areas, known as the drylands, and has a particular focus on Africa. As of October 2018, the UNCCD had 197 Parties.

United Nations Framework Convention on Climate Change (UNFCCC)

The UNFCCC was adopted in May 1992 and opened for signature at the 1992 Earth Summit in Rio de Janeiro. It entered into force in March 1994 and as of May 2018 had 197 Parties (196 States and the European Union). The Convention's ultimate objective is the 'stabilisation of greenhouse gas concentrations in the atmosphere at a level that would prevent dangerous anthropogenic interference with the climate system'. The provisions of the Convention are pursued and implemented by two treaties: the *Kyoto Protocol* and the *Paris Agreement*.

Urban green infrastructure

Public and private green spaces, including remnant native vegetation, parks, private gardens, golf courses, street trees, urban farming and engineered options such as green roofs, green walls, biofilters and raingardens (Norton et al., 2015).

Urban and Peri-urban agriculture

'The cultivation of crops and rearing of animals for food and other uses within and surrounding the boundaries of cities, including fisheries and forestry' (EPRS, 2014).

Uptake

The addition of a substance of concern to a *reservoir*.

See also *Carbon sequestration*, and *Sink*.

Vulnerability

The propensity or predisposition to be adversely affected. Vulnerability encompasses a variety of concepts and elements including sensitivity or susceptibility to harm and lack of capacity to cope and adapt.

See also *Exposure*, *Hazard*, and *Risk*.

Water cycle

See *Hydrological cycle*.

Wellbeing

A state of existence that fulfils various human needs, including material living conditions and quality of life, as well as the ability to pursue one's goals, to thrive, and feel satisfied with one's life. Ecosystem well-being refers to the ability of *ecosystems* to maintain their diversity and quality.

Wetland

Land that is covered or saturated by water for all or part of the year (e.g., *peatland*).

References

United Nations Secretary General's Advisory Group on Energy and Climate (AGECC), 2010: *Energy for a Sustainable Future*. New York, NY, USA.

Allan, J.A., 2005: Virtual water: A strategic resource global solutions to regional deficits. *Groundwater*, 36(4), 545-546, doi: [10.1111/j.1745-6584.1998.tb02825.x](https://doi.org/10.1111/j.1745-6584.1998.tb02825.x).

AMS, 2000: AMS Glossary of Meteorology, 2nd ed. American Meteorological Society, Boston, MA, Retrieved from: <http://amsglossary.allenpress.com/glossary/browse>.

Arctic Council, 2013: Glossary of terms. In: Arctic Resilience Interim Report 2013. Stockholm Environment Institute and Stockholm Resilience Centre, Stockholm, Sweden, pp. viii.

Arctic Council, 2016: Arctic Resilience Report 2016. , eds. M. Carson and G. Peterson Stockholm Sweden: Stockholm Environment Institute and Stockholm Resilience Centre.

Berkes, F. and C. Folke, 1998: *Linking Social and Ecological Systems: Management Practices and Social Mechanisms for Building Resilience*. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 459 pp.

CGIAR, 2019: Sustainable intensification of agriculture: oxymoron or real deal? The Consultative Group on International Agricultural Research (CGIAR). Retrieved from: <https://wle.cgiar.org/thrive/big-questions/sustainable-intensification-agriculture-oxymoron-or-real-deal/sustainable-1> .

Culwick, C. and K. Bobbins, 2016: A Framework for a Green Infrastructure Planning Approach in the Gauteng City–Region. GCRO Research Report No. 04, Gauteng City–Region Observatory (GRCO), Johannesburg, South Africa, 127 pp.

Davis, S.J. and K. Caldeira, 2010: Consumption-based accounting of CO₂ emissions. *Proceedings of the National Academy of Sciences of the United States of America*, 107(12), 5687-5692, doi:[10.1073/pnas.0906974107](https://doi.org/10.1073/pnas.0906974107).

Dietary Guidelines Advisory Committee, 2015: Scientific Report of the 2015 Dietary Guidelines Advisory Committee. US Department of Agriculture, Agricultural Research Service, Washington DC, United States of America. Retrieved from: https://ods.od.nih.gov/pubs/2015_dgac_scientific_report.pdf.

Ellis, E.C. and N. Ramankutty, 2008: Putting people in the map: anthropogenic biomes of the world. *Front Ecol Environ* 6(8), 439-447, doi: [10.1890/070062](https://doi.org/10.1890/070062).

Ellis, E.C., K.K. Goldewijk, S. Siebert, D. Lightman and N. Ramankutty, 2010: Anthropogenic transformation of the biomes, 1700 to 2000. *Global Ecology and Biogeography*, 19(5), 589-606, doi:[10.1111/j.1466-8238.2010.00540.x](https://doi.org/10.1111/j.1466-8238.2010.00540.x).

EPRS, 2014: Urban and Peri-Urban Agriculture. European Parliamentary Research Service. Retrieved from: <https://epthinktank.eu/2014/06/18/urban-and-peri-urban-agriculture/>

FAO, 1985: *Irrigation water management: Trianing manual no.1 – Introduction to irrigation*. Food and edws, C. Brouwer, A. Goffeau, and M. Heibloem. Agriculture Organisation of the United Nations (FAO), Rome, Italy.

FAO, 2000: Land cover classification system (LCCS): Classification concepts and user manual. eds. A. Di Gregorio and L.J.M. Jansen. Food and Agriculture Organisation of the United Nations (FAO), Rome, Italy, 179 pp. Retrieved from: <http://www.fao.org/docrep/003/x0596e/x0596e01e.htm>.

FAO, 2001: Glossary. In: The State of Food Insecurity in the World 2001. Food and Agriculture Organisation of the United Nations (FAO), Rome, Italy, pp. 49–50.

FAO, 2005: Building on gender, agrobiodiversity and local knowledge: A training manual. Food and Agriculture Organisation of the United Nations (FAO), Rome, Italy.

FAO, 2007: Land evaluation: Towards a revised framework. Land and water discussion paper. Food and Agriculture Organisation of the United Nations (FAO), Rome, Italy.

FAO, 2009: Declaration of the World Summit on Food Security. WSFS 2009/2, Food and Agriculture Organisation of the United Nations (FAO), Rome, Italy.

FAO, 2013: Food wastage footprint: Impacts on natural resources. Summary report. Food and Agriculture Organization of the United Nations (FAO), Rome, Italy, 63 pp.

FAO, 2015a: Agroforestry. Food and Agriculture Organization of the United Nations (FAO). Retrieved from: <http://www.fao.org/forestry/agroforestry/80338/en/>.

FAO, 2015b: Food waste. Food and Agriculture Organization of the United Nations (FAO). Retrieved from: <http://www.fao.org/platform-food-loss-waste/food-waste/definition/en/>.

FAO, 2018: Climate-Smart Agriculture. Food and Agriculture Organization of the United Nations (FAO). Retrieved from: www.fao.org/climate-smart-agriculture.

Forest Europe, 1993: Resolution H1 General Guidelines for the Sustainable Management of Forests in Europe. Second Ministerial Conference on the Protection of Forests in Europe, Helsinki, 16-17 June 1993. Retrieved from: https://www.foresteurope.org/docs/MC/MC_helsinki_resolutionH1.pdf.

Fung, A. and E.O. Wright (eds.), 2003: Deepening Democracy: Institutional Innovations in Empowered Participatory Governance. Verso, London, UK, 312 pp.

Garnett, T., M.C. Appleby, A. Balmford, I.J. Bateman, T.G. Benton, P. Bloomer, B. Burlingame, M. Dawkins, L. Dolan, D. Fraser, M. Herrero, I. Hoffman, P. Smith, P.K. Thornton, C. Toulmin, S.J. Vermeulen, H.C.J. Godfray, 2013: Sustainable intensification in Agriculture: Premises and Policies. *Science*. 341, 6141, 33. doi:10.1126/science.1234485.

Gbeckor-Kove, N. 1989: Lectures on drought, desertification, in drought and desertification. WMO, TDNo. 286, World Meteorological Organization, Geneva, Switzerland.

HLPE, 2017: Nutrition and food systems. A report by the High Level Panel of Experts on Food Security and Nutrition of the Committee on World Food Security, Rome. Retrieved from: <http://www.fao.org/3/a-i7846e.pdf>.

IBI, 2018: Frequently Asked Questions About Biochar: What is biochar? International Biochar Initiative (IBI). Retrieved from: <https://biochar-international.org/faqs>.

ICS, 2019: Formal subdivision of the Holocene Series/Epoch. International Commission on Stratigraphy (ICS). Retrieved from <http://www.stratigraphy.org/index.php/ics-news-and-meetings/125-formal-subdivision-of-the-holocene-series-epoch>.

IOM, 2018: Key Migration Terms. International Organization for Migration (IOM). Retrieved from: www.iom.int/key-migration-terms.

IPBES, 2018: The IPBES assessment report on land degradation and restoration. [Montanarella, L., Scholes, R., and Brainich, A. (eds.)]. Secretariat of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem services, Bonn, Germany, 744 pp.

IPBES, 2019: Glossary. Secretariat of the Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem services, Bonn, Germany. Retrieved from: <https://www.ipbes.net/glossary>.

IPCC, 2000: Land Use, Land-Use Change, and Forestry: A Special Report of the IPCC. [Watson, R.T., I.R. Noble, B. Bolin, N.H. Ravindranath, D.J. Verardo, and D.J. Dokken (eds.)]. Cambridge University Press, Cambridge, UK, 375 pp.

IPCC, 2003: Definitions and Methodological Options to Inventory Emissions from Direct Human-induced Degradation of Forests and Devegetation of Other Vegetation Types. [Penman, J., M. Gytarsky, T. Hiraishi, T. Krug, D. Kruger, R. Pipatti, L. Buendia, K. Miwa, T. Ngara, K. Tanabe, and F. Wagner (eds.)]. Institute for Global Environmental Strategies (IGES), Hayama, Kanagawa, Japan, 32 pp.

IPCC, 2004: IPCC Workshop on Describing Scientific Uncertainties in Climate Change to Support Analysis of Risk of Options. Workshop Report. Intergovernmental Panel on Climate Change (IPCC), Geneva, Switzerland, 138 pp.

IPCC, 2006: IPCC Guidelines for National Greenhouse Gas Inventories. [H.S. Eggleston, L. Buendia, K. Miwa, T. Ngara, K. Tanabe (eds.)]. Institute for Global Environmental Strategies (IGES), Hayama, Kanagawa, Japan, 20 pp.

IPCC, 2011: Workshop Report of the Intergovernmental Panel on Climate Change Workshop on Impacts of Ocean Acidification on Marine Biology and Ecosystems. [Field, C.B., V. Barros, T.F. Stocker, D. Qin, K.J. Mach, G.-K. Plattner, M.D. Mastrandrea, M. Tignor, and K.L. Ebi (eds.)]. IPCC Working Group II Technical Support Unit, Carnegie Institution, Stanford, California, United States of America, 164 pp.

IPCC, 2012a: Managing the Risks of Extreme Events and Disasters to Advance Climate Change Adaptation. A Special Report of Working Groups I and II of the Intergovernmental Panel on Climate Change (IPCC). [Field, C.B., V. Barros, T.F. Stocker, D. Qin, D.J. Dokken, K.L. Ebi, M.D. Mastrandrea, K.J. Mach, G.-K. Plattner, S.K. Allen, M. Tignor, and P.M. Midgley (eds.)]. Cambridge University Press, Cambridge, UK and New York, NY, USA, 582 pp.

IPCC, 2012b: Meeting Report of the Intergovernmental Panel on Climate Change Expert Meeting on Geoengineering. IPCC Working Group III Technical Support Unit, Potsdam Institute for Climate Impact Research, Potsdam, Germany, 99 pp.

IPCC, 2013: 2013 Supplement to the 2006 IPCC Guidelines for National Greenhouse Gas Inventories: Wetlands. [Hiraishi, T., T. Krug, K. Tanabe, N. Srivastava, J. Baasansuren, M. Fukuda, and T.G. Troxler (eds.)]. IPCC, Switzerland.

ISO, 2014: ISO 16559:2014(en) Solid biofuels – Terminology, definitions and descriptions. International Standards Organisation (ISO). Retrieved from: <https://www.iso.org/obp/ui/#iso:std:iso:16559:ed-1:v1:en>.

ISO, 2018: ISO 14044:2006. Environmental management – Life cycle assessment – Requirements and guidelines. International Standards Organisation (ISO). Retrieved from: www.iso.org/standard/38498.html.

Jagers, S.C. and J. Stripple, 2003: Climate Governance Beyond the State. *Global Governance*, **9(3)**, 385–399, www.jstor.org/stable/27800489.

MacDonald, G.K., K.A. Brauman, S. Sun, K.M. Carlson, E.S. Cassidy, J.S. Gerber, and P.C. West, 2015: Rethinking agricultural trade relationships in an era of globalization. *BioScience* 65(3), 275–289, doi:[10.1093/biosci/biu225](https://doi.org/10.1093/biosci/biu225).

Mastrandrea, M.D. et al., 2010: Guidance Note for Lead Authors of the IPCC Fifth Assessment Report on Consistent Treatment of Uncertainties. Intergovernmental Panel on Climate Change (IPCC), Geneva, Switzerland, 6 pp.

McDonald, T., J. Jonson, and K.W. Dixon, 2016: National standards for the practice of ecological restoration in Australia. *Restoration Ecology*, 24(S1) S4–S32, doi:[10.1111/rec.12359](https://doi.org/10.1111/rec.12359).

MA, 2005: Appendix D: Glossary. In: *Ecosystems and Human Well-being: Current States and Trends. Findings of the Condition and Trends Working Group* [Hassan, R., R. Scholes, and N. Ash (eds.)]. Millennium Ecosystem Assessment (MA). Island Press, Washington DC, USA, pp. 893–900.

Meinzen-Dick, R. and M. Di Gregorio, 2004: Collective Action and Property Rights for Sustainable Development. International Food and Policy Research Institute. Washington DC, USA.

Mitchell, T. and S. Maxwell, 2010: Defining climate compatible development. CDKN ODI Policy Brief November 2010/A, Climate & Development Knowledge Network (CDKN), 6 pp.

Misra, A., L. Khurana, 2008: Obesity and the metabolic syndrome in developing countries. *The Journal of Clinical Endocrinology & Metabolism* 93(11), 3–30, doi: [10.1210/jc.2008-1595](https://doi.org/10.1210/jc.2008-1595).

Moss, R.H. and S.H. Schneider, 2000: Uncertainties in the IPCC TAR: Recommendations to Lead Authors for More Consistent Assessment and Reporting. In: *Guidance Papers on the Cross Cutting Issues of the Third Assessment Report of the IPCC* [Pachauri, R., T. Taniguchi, and K. Tanaka (eds.)]. Intergovernmental Panel on Climate Change (IPCC), Geneva, Switzerland, pp. 33–51

Moss, R.H. et al., 2008: Towards New Scenarios for Analysis of Emissions, Climate Change, Impacts, and Response Strategies. Technical Summary. Intergovernmental Panel on Climate Change (IPCC), Geneva, Switzerland, 25 pp.

Moss, R.H. et al., 2010: The next generation of scenarios for climate change research and assessment. *Nature*, **463(7282)**, 747–756, doi:[10.1038/nature08823](https://doi.org/10.1038/nature08823).

MRFCJ, 2018: Principles of Climate Justice. Mary Robinson Foundation For Climate Justice (MRFCJ). Retrieved from: www.mrfcj.org/principles-of-climate-justice.

NDSU, 2014: Saline and Sodic Soils. North Dakota State University (NDSU). Retrieved from: <https://www.ndsu.edu/soilhealth/wp-content/uploads/2014/07/Saline-and-Sodic-Soils-2-2.pdf>.

Norton, B.A., A.M. Coutts, S.J. Livesley, R.J. Harris, A.M. Hunter, and N.S.G. Williams, 2015: Planning for cooler cities: A framework to prioritise green infrastructure to mitigate high temperatures in urban landscapes. *Landscape and Urban Planning*, 134, 127–138, doi:[10.1016/j.landurbplan.2014.10.018](https://doi.org/10.1016/j.landurbplan.2014.10.018).

O'Neill, B.C. et al., 2014: A new scenario framework for climate change research: the concept of shared socioeconomic pathways. *Climatic Change*, **122(3)**, 387–400, doi:[10.1007/s10584-013-0905-2](https://doi.org/10.1007/s10584-013-0905-2).

O'Neill, B.C. et al., 2017: The roads ahead: Narratives for shared socioeconomic pathways describing world futures in the 21st century. *Global Environmental Change*, **42**, 169–180, doi:[10.1016/j.gloenvcha.2015.01.004](https://doi.org/10.1016/j.gloenvcha.2015.01.004).

Park, S. E., N.A. Marshall, E. Jakku, A.M Dowd, S.M. Howden, E.K. Mendham and A. Fleming, 2012: Informing adaptation responses to climate change through theories of transformation. *Global Environmental Change*, 22(1), 115–126. doi:[10.1016/j.gloenvcha.2011.10.003](https://doi.org/10.1016/j.gloenvcha.2011.10.003).

Peters, B.G. and J. Pierre, 2001: Developments in intergovernmental relations: towards multi-level governance. *Policy & Politics*, **29(2)**, 131–135, doi:[10.1332/0305573012501251](https://doi.org/10.1332/0305573012501251).

Popkin, B.M., 1994: The nutrition transition in low-income countries: an emerging crisis. *Nutr Rev* 52(9), 285-298, doi: <https://doi.org/10.1111/j.1753-4887.1994.tb01460.x>.

Popkin, B.M., 2006: Global nutrition dynamics: the world is shifting rapidly toward a diet linked with noncommunicable diseases. *The American Journal of Clinical Nutrition*, 84(2), 289-298, doi: <https://doi.org/10.1093/ajcn/84.1.289>.

Pongratz, J. et al., 2018: Models meet data: Challenges and opportunities in implementing land management in Earth system models. *Global Change Biology*, 24(4) 1470-1487, doi: [10.1111/gcb.13988](https://doi.org/10.1111/gcb.13988).

Riahi, K. et al., 2017: The Shared Socioeconomic Pathways and their energy, land use, and greenhouse gas emissions implications: An overview. *Global Environmental Change*, **42**, 153–168, doi:[10.1016/j.gloenvcha.2016.05.009](https://doi.org/10.1016/j.gloenvcha.2016.05.009).

Rulli, M.C. and P. D'Odorico, 2014: Food appropriation through large scale land acquisitions. *Environmental Research Letters*, 9(6), doi: [10.1088/1748-9326/9/6/064030](https://doi.org/10.1088/1748-9326/9/6/064030).

Sarmiento, H. and C. Tilly, 2018: Governance Lessons from Urban Informality. *Politics and Governance*, 6(1), 199–202, doi:[10.17645/pag.v6i1.1169](https://doi.org/10.17645/pag.v6i1.1169).

Silvertown, J. (2009). A new dawn for citizen science. *Trends in Ecology & Evolution*, 24(9), 467-471, doi: [10.1016/J.TREE.2009.03.017](https://doi.org/10.1016/J.TREE.2009.03.017).

Soulé, M.E., 2010: Conservation relevance of ecological cascade. In: *Trophic Cascades: Predators, Prey, and the Changing Dynamics of Nature*. eds. J. Terborgh and J.A. Estes. Island Press, 337-352.

Tabara, J.D., J. Jager, D. Mangalagiu, and M. Grasso, 2018: Defining transformative climate science to address high-end climate change. *Regional Environmental Change*, 1–12, doi:[10.1007/s10113-018-1288-8](https://doi.org/10.1007/s10113-018-1288-8).

Termeer, C.J.A.M., A. Dewulf, and G.R. Biesbroek, 2017: Transformational change: governance interventions for climate change adaptation from a continuous change perspective. *Journal of Environmental Planning and Management*, **60(4)**, 558–576, doi:[10.1080/09640568.2016.1168288](https://doi.org/10.1080/09640568.2016.1168288).

Türkeş M, 1999: Vulnerability of Turkey to desertification with respect to precipitation and aridity conditions. *Turkish J. Eng. Environ. Sci.*, 23, 363-380.

Türkeş, M., 2017: *General Climatology: Fundamentals of Atmosphere, Weather and Climate*. Revised Second Edition, Kriter Publisher Physical Geography Series No: 4, ISBN: 978-605-9336-28-4, xxiv + 520 pp. Kriter Publisher, Berdan Matbaası: İstanbul. (In Turkish).

UN, 1992: Article 2: Use of Terms. In: Convention on Biological Diversity. United Nations (UN), pp. 3–4.

UN, 1998: Guiding Principles on Internal Displacement. E/CN.4/1998/53/Add.2, United Nations (UN) Economic and Social Council, 14 pp.

UN, 2015: Transforming Our World: The 2030 Agenda for Sustainable Development. A/RES/70/1, United Nations General Assembly (UNGA), New York, NY, USA, 35 pp.

UNCCD, 1994: United Nations Convention to Combat Desertification in countries experiencing serious drought and/or desertification, particularly in Africa. A/AC.241/27, United Nations General Assembly (UNGA), New York, NY, USA, 58 pp.

UNCCD, 2019: Achieving Land Degradation Neutrality. United Nations Convention to Combat Desertification. Retrieved from: <https://www.unccd.int/actions/achieving-land-degradation-neutrality>.

UN DESA, 2016: Identifying social inclusion and exclusion. In: Leaving no one behind: the imperative of inclusive development. Report on the World Social Situation 2016. ST/ESA/362, United Nations Department of Economic and Social Affairs (UN DESA), New York, NY, USA, pp. 17–31.

UNESCO, 2018: Local and Indigenous Knowledge Systems. United Nations Educational, Scientific and Cultural Organization (UNESCO). Retrieved from: www.unesco.org/new/en/natural-sciences/priority-areas/links/related-information/what-is-local-and-indigenous-knowledge.

UNEP, 2016: Unlocking the sustainable potential of land resources: evaluating systems, strategies and tools. Factsheet. United Nations Environment Programme (UNEP). Retrieved from: <http://hdl.handle.net/20.500.11822/7711>.

UNFCCC, 1992: United Nations Framework Convention on Climate Change. United Nations, New York, United States. Retrieved from: https://unfccc.int/files/essential_background/background_publications_htmlpdf/application/pdf/conve ng.pdf.

UNFCCC, 2019: Land Use, Land-Use Change and Forestry (LULUCF). United Nations Framework Convention on Climatic Change (UNFCCC), Bonn, Germany. Retrieved from: <https://unfccc.int/topics/land-use/workstreams/land-use--land-use-change-and-forestry-lulucf>.

UNISDR, 2009: 2009 UNISDR Terminology on Disaster Risk Reduction. United Nations International Strategy for Disaster Reduction (UNISDR), Geneva, Switzerland, 30 pp.

UNOHCHR, 2018: What are Human rights? UN Office of the High Commissioner for Human Rights (UNOHCHR). Retrieved from: www.ohchr.org/EN/Issues/Pages/whatarehumanrights.aspx.

UN–OHRLLS, 2011: Small Island Developing States: Small Islands Big(ger) Stakes. Office for the High Representative for the Least Developed Countries, Landlocked Developing Countries and Small Island Developing States (UN–OHRLLS), New York, NY, USA, 32 pp.

UN–OHRLLS, 2018: Small Island Developing States: Country profiles. Office for the High Representative for the Least Developed Countries, Landlocked Developing Countries and Small Island Developing States (UN–OHRLLS). Retrieved from: <http://unohrlls.org/about-sids/country-profiles>.

UN-REDD, 2009: Measurement, Assessment, Reporting and Verification (MARV): Issues and Options for REDD. Draft Discussion Paper, United Nations Collaborative Programme on Reducing Emissions from Deforestation and Forest Degradation in Developing Countries (UN-REDD), Geneva, Switzerland, 12 pp.

WCED, 1987: Our Common Future. World Commission on Environment and Development (WCED), Geneva, Switzerland, 400 pp., doi:[10.2307/2621529](https://doi.org/10.2307/2621529).

Vandergeten, E., H. Azadi, D. Teklemariam, J. Nyssen, F. Witlox and E. Vanhaute, 2016: Agricultural Outsourcing or Land Grabbing: A Meta-Analysis. *Landscape Ecology*, 31(7), 1395-1417, doi: [10.1007/s10980-016-0365-y](https://doi.org/10.1007/s10980-016-0365-y).

Well, R.R. and N.C. Brady, 2016: Nature and property of soils, 15th Edition. Pearson Education Limited, Harlow, England, 420-430 pp.

Wiener, J.B. and J.D. Graham, 2009: Risk vs. risk: Tradeoffs in protecting health and the environment. Harvard University Press, Cambridge, MA, USA.

Willems, S. and K. Baumert, 2003: Institutional Capacity and Climate Actions. COM/ENV/EPOC/IEA/SLT(2003)5, Organisation for Economic Co-operation and Development (OECD) International Energy Agency (IEA), Paris, France, 50 pp.

Whitmee, S. et al., 2015: Safeguarding human health in the Anthropocene epoch: Report of the Rockefeller Foundation-Lancet Commission on planetary health. *The Lancet*, 366(10007), 1973-2028.

WHO, 2018: Malnutrition. World Health Organization (WHO). Retrieved from: <http://www.who.int/topics/malnutrition/en/>.

WOCAT, undated: Glossary. World Overview of Conservation Approaches and Technologies (WOCAT). Retrieved from: <https://www.wocat.net/en/glossary>.

WMO, 2019: What are weather/climate services? Global Framework for Climate Services, Geneva, Switzerland. Retrieved from: https://www.wmo.int/gfcs/what_are_climate_weather_services.

Young, A. 1989: Agroforestry for soil conservation. International Council for Research in Agroforestry, Nairobi, Kenya, 318p.

From Edvin Aldrian, IPCC Working Group I Vice Chair and Review Editor
Chapter One of the SRCCL

Prepared in collaboration with María José Sanz and Bruce McCarl

Re Review editor report on Chapter 1 in **Climate Change and Land: an IPCC Special Report on climate change, desertification, land degradation, sustainable land management, food security and greenhouse gas fluxes in terrestrial ecosystems**

Here we note activities undertaken by myself and the other chapter 1 review editors plus our opinions on the author team responses and interactions.

First before interacting with the author team I responded to a TSU email providing names of some suggested reviewers. I also noted during the response process that several of the suggested people did provide comments.

Second, myself and the other chapter 1 review editors worked with the Chapter 1 author team in responding to comments during the our attendance at the 3rd and 4th authors meetings. In working with the author team we as review editors: a) carefully read the FOD and SOD chapter drafts in preparation for a meeting with the author team over the review comments; b) considered each comment in response to the FOD and SOD drafts identifying major classes of comments and a judgement on priority of a response to the classes and comments therein; c) made a preliminary list of cross-cutting issues raised in the comments that might be best left to other chapters or might require a joint response coordinated with other chapter; d) dialogued with the author team on ways to respond to the comments including developing a coded system of basic responses; e) discussed the role of an introductory chapter as opposed to more in depth chapters; f) made suggestions on the fundamental nature of author team redrafting to address reviewer comments; g) made suggestions on issues to add and sections to shorten in both response to comments and in an effort to better fit the chapter into the total SRCCL document as an introductory chapter; h) did not do any of the redrafting and content alteration in the chapter; i) evaluated all of the author team responses and any associated redrafting done in reaction to both FOD and SOD draft stimulated comments; j) in cases interacted with the author team during the drafting period identifying needs for more complete responses and then evaluated any response or drafting alterations and k) refrained from commenting in the formal process on the FOD or SOD.

In working through this process we noted several significant differences of opinion between the team and the commenters that required resolution. The notable points of disagreement involved a) chapter role as an introduction and b) appropriate content to include in an introduction. In particular a number of commenters criticized what we considered to be excess detail in the chapter that rather could rely on treatments in the rest of the SRCCL or called for detail beyond what the chapter had available pages and scope to cover. In those cases we dialogued over increased use of forward references in the SRCCL, more references to fundamental treatments elsewhere in the literature and an author evaluation process that considered whether some detailed draft sections really belonged in the document introduction. As a consequence we saw a strong reorientation in the revision from more of a subject matter chapter to more of an introductory chapter.

In between authors meetings we received several versions of the responses and associated changes in the draft from the Chapter 1 team members or the chapter scientist and held an iterative discussion on how to respond to the comments and implicit criticisms. The result was in cases a change in the chapter text and in other cases no changes were

made. Nevertheless in all cases the team constructed a satisfactory response that we felt was quite adequate and addressed the comment nicely. In our opinion, the comments lessened in severity for the SOD relative to the FOD indicating a good response from the whole team during the FOD process. SOD responses were also good.

The result of this was an appropriate chapter 1 revision in reaction to both FOD and SOD comments that took into account the vast majority of the writing, referencing, and subject matter comments. This was coupled with responses to a subset of the comments that referred the commenters to more complete treatments in other chapters or responses indicating some comments were beyond the possible scope of the introductory chapter.

In my judgement and that of my fellow review editors we feel all review comments have been sufficiently addressed in the responses or redrafting.

From Bruce A. McCarl, University Distinguished Professor of Agricultural Economics, Texas A&M University and Review Editor Chapter One of the SRCCL

Prepared in collaboration with María José Sanz and Edwin Aldrian

Re Review editor report on Chapter 1 in **Climate Change and Land: an IPCC Special Report on climate change, desertification, land degradation, sustainable land management, food security and greenhouse gas fluxes in terrestrial ecosystems**

Date July 5, 2019

Here we note activities undertaken by myself and the other chapter 1 review editors plus our opinions on the author team responses and interactions.

First before interacting with the author team I responded to a TSU email providing names of some suggested reviewers. I also noted during the response process that several of the suggested people did provide comments.

Second, myself and the other chapter 1 review editors worked with the Chapter 1 author team in responding to comments during the our attendance at the 3rd and 4th authors meetings. In working with the author team we as review editors: a) carefully read the FOD and SOD chapter drafts in preparation for a meeting with the author team over the review comments; b) considered each comment in response to the FOD and SOD drafts identifying major classes of comments and a judgement on priority of a response to the classes and comments therein; c) made a preliminary list of cross-cutting issues raised in the comments that might be best left to other chapters or might require a joint response coordinated with other chapter; d) dialogued with the author team on ways to respond to the comments including developing a coded system of basic responses; e) discussed the role of an introductory chapter as opposed to more in depth chapters; f) made suggestions on the fundamental nature of author team redrafting to address reviewer comments; g) made suggestions on issues to add and sections to shorten in both response to comments and in an effort to better fit the chapter into the total SRCCL document as an introductory chapter; h) did not do any of the redrafting and content alteration in the chapter; i) evaluated all of the author team responses and any associated redrafting done in reaction to both FOD and SOD draft stimulated comments; j) in cases interacted with the author team during the drafting period identifying needs for more complete responses and then evaluated any response or drafting alterations and k) refrained from commenting in the formal process on the FOD or SOD.

In working through this process we noted several significant differences of opinion between the team and the commenters that required resolution. The notable points of disagreement involved a) chapter role as an introduction and b) appropriate content to include in an introduction. In

particular a number of commenters criticized what we considered to be excess detail in the chapter that rather could rely on treatments in the rest of the SRCCL or called for detail beyond what the chapter had available pages and scope to cover. In those cases we dialogued over increased use of forward references in the SRCCL, more references to fundamental treatments elsewhere in the literature and an author evaluation process that considered whether some detailed draft sections really belonged in the document introduction. As a consequence we saw a strong reorientation in the revision from more of a subject matter chapter to more of an introductory chapter.

In between authors meetings we received several versions of the responses and associated changes in the draft from the Chapter 1 team members or the chapter scientist and held an iterative discussion on how to respond to the comments and implicit criticisms. The result was in cases a change in the chapter text and in other cases no changes were made. Nevertheless in all cases the team constructed a satisfactory response that we felt was quite adequate and addressed the comment nicely. In our opinion, the comments lessened in severity for the SOD relative to the FOD indicating a good response from the whole team during the FOD process. SOD responses were also good.

The result of this was an appropriate chapter 1 revision in reaction to both FOD and SOD comments that took into account the vast majority of the writing, referencing, and subject matter comments. This was coupled with responses to a subset of the comments that referred the commenters to more complete treatments in other chapters or responses indicating some comments were beyond the possible scope of the introductory chapter.

In my judgement and that of my fellow review editors we feel all review comments have been sufficiently addressed in the responses or redrafting.

From María José Sanz, Basque Centre for Climate Change (BC3) and Review Editor
Chapter One of the SRCCL

Prepared in collaboration with Bruce A. McCarl and Edvin Aldrian

Re Review editor report on Chapter 1 in **Climate Change and Land: an IPCC Special Report on climate change, desertification, land degradation, sustainable land management, food security and greenhouse gas fluxes in terrestrial ecosystems**

Date July 10, 2019

As review editors we interacted with followed the actions taken by the authors to address the comments given by experts and governments, here we summarize our recollection of the process.

Myself and the other chapter 1 review editors worked with the authors team in responding to comments by experts and governments during and between our attendance at the 3rd and 4th Lead Author meetings.

In working with the author team we as review editors: a) after considering the FOD and SOD chapter drafts, went over the all review comments and got an overview of the issues raised, as well as identified cluster of comments dealing with similar topic/issue; b) shared our views on the cluster of comments and made a preliminary list of issues raised in the comments that might be best left to other chapters or might require a joint response coordinated with other chapter; c) dialogued with the author team on ways to respond to the comments including developing a coded system of basic responses; d) given the fact that many comments raised the issue of the role of an introductory chapter as opposed to more in depth chapters, we discussed with the authors how to address this important overall framing comment; e) made suggestions on issues to add and sections to shorten in both response to comments and in an effort to better fit the chapter into the total SRCCL document as a framing chapter, yet to keep in it the necessary content; f) did not do any of the redrafting and content alteration in the chapter; g) evaluated all of the author team responses and any associated redrafting done in reaction to both FOD and SOD draft stimulated comments; h) interacted with the author team during the drafting period identifying needs for more complete responses and then evaluated any response or drafting alterations, this was done in several iterations to ensure consistency with the new drafts and completeness of the responses; and i) refrained from commenting in the formal process on the FOD or SOD.

In working through this process notable points of disagreement respect to the comments were expressed by the authors that related mostly to issue of the extent to which the chapter should include detailed information or be more and introduction to the report. The fact that the chapter was disconnected from the rest of the chapters and that in some cases contradicting the content of other chapters, including in the treatment or inclusion of references. As a

consequence, authors made a great effort in reorientation in the revision, and in the final draft the chapter was better frame and reflected the structure and provided good hints for navigating the rest of the report.

In our opinion, the comments lessened in severity for the SOD relative to the FOD indicating reflecting the whole team efforts in addressing the comments during the FOD process.

The result of this was an appropriate chapter 1 revision in reaction to both FOD and SOD comments that took into account the vast majority of the writing, referencing, and subject matter comments. This was coupled with responses to a subset of the comments that referred the commenters to more complete treatments in other chapters or responses indicating some comments were beyond the possible scope of the introductory chapter.

In my judgement and that of my fellow review editors we feel all review comments have been sufficiently addressed in the responses by the authors.

Final Review Editor Report

IPCC SRCCL

Chapter 2 « Land-Climate Interactions »

Pierre Bernier

1. INTRODUCTION:

This final report follows two interim report by all three review editors assigned to Chapter 2. Unlike the interim report, this last final report is to be submitted individually by each review editor, and is meant to identify outstanding issues. In particular, in this short text, I will therefore cover the following points:

- Issues with the review process
- Potential controversies or disagreements in the text
- Other general remarks

2. ISSUES WITH THE REVIEW PROCESS

This was my first time as a participant to an IPCC report process, not counting my participations as expert reviewer over the years. I entered my RE role with only a theoretical understanding of my responsibilities. My first observation at LAM 3 was that the role of Review Editor was officially defined narrowly (to evaluate the responses of authors to comments by reviewers), but that unwritten expectations were much broader. In fact, my impression was that an RE was expected to do all in his or her capacity to support the production of a high quality document. This included at a minimum being very pro-active in getting authors to consider comments and provide responses, but could also include helping clarify or correct text, up to providing content if necessary.

The result of this difference between the defined and the expected roles was that all REs were not on the same page in this how to approach their tasks. Within our small group, attitudes ranged from simply evaluating authors responses as provided, to being very hands-on and pro-active across all aspects of document development and review. Because of the number of comments (nearly 3000), we each took on the responsibility for different portions of the chapter. But because of the range of attitudes, all portions of the text did not get the same level of support and scrutiny by review editors.

Given this issue, I would recommend that the TSU provide a better description of its expectations with respect to the role of the Review Editors. With this in hand, review editors for a given chapter could agree among themselves to a common standard with respect to their role.

There was also an issue of timeline. I personally I tried as much as possible to use the review comments to help authors improve their text. However, for this I needed to get author responses before final text submission to the TSU so that there was time to give feedback to the authors. I believed that this was particularly critical for the SOD because this was the last chance at correction. Given the very tight deadline, I

was only partially successful in this task as we received a large fraction of the responses to comments after the text submission deadline to the TSU.

3. POTENTIAL CONTROVERSIES OR DISAGREEMENTS IN THE TEXT

There were no controversies among authors as to specific technical points within the text, in part because there was not a great deal of overlap in expertise between the different authors given that large breadth of topics covered.

The only real controversy that came up was with respect to the organization of the chapter itself. Following LAM 4 and substantial discussions, the team decided to radically modify a section of the text, dispatching portions to other sections, and re-purposing the remainder by moving it to the end of the chapter as supporting information to the previous section. The resulting changes to all section numbers, two weeks before submission deadline to the TSU, had all sorts of ramifications including notifying authors in other chapters so that references to our chapter would use the new section numbers. This should have been done much earlier in the process, but lack of agreement on other points as well as personality issues delayed this decision.

In terms of controversy on scientific points, I believe that the portion of section 6 (former section 7) dealing with results from Integrated Assessment Models (IAMs) may raise some thorny concerns with some readers. As mentioned in the second interim report, reviewers raised two issues with the IAM models. The first was that these models do not represent biophysical feedbacks such as albedo in their modelling structure. The lack of biophysical feedback was striking since other sections of chapter 2 specifically state their feedbacks. In fact, one of the important points raised by the chapter is that albedo feedback can overwhelm the carbon-related effect of afforestation in boreal forests and thus may reduce the desirability of current or proposed policies of increasing forest-based carbon sequestration in these regions. Not having albedo in IAM models may blur this message.

The second issue with IAMs that was raised by the reviewers is that they do not incorporate socio-economic or other feasibility constraints. This turned out to be particularly critical to reviewers, including Debra Roberts in her FOD review, for BECCS, one of the preferred mitigation options in many of the IAM outputs. In fact, BECCS as a large-scale mitigation option faces two hurdles: the competition for land between large-scale biomass production and food production, and yet-to-be demonstrate scalability of CCS to the industrial-scale. In response to these comments, the authors only put a mild caveat statement mostly pointing to another chapter for further explanations.

4. OTHER GENERAL REMARKS

I first want to highlight the efforts made by Elena Shevliakova as the “lead” CLA to get the agenda moving forward. This was not an easy task as authors were not all equally cooperative during meetings or responsive to emails once the LAM was over. I also want to highlight the role played with great efficiency by Nathalie de Noblet Ducoudré who very took over the lead of LAM4 in replacement of Elena absent due to the US government shutdown. Natalie succeeded in getting most of the overlap among sections resolved, a process that involved outstanding collaboration between authors on the final day of LAM4.

On the negative side, I cannot keep silent on the near-absolute non-participation of Andrey Sirin who did not say a single word during the total length of LAM4 and barely participated in email exchanges. I would also like to point out that we had at least one otherwise excellent author that was also involved in AR6 and that failed to provide responses to all review comments. I suspect that her overlap in commitments caused her disengagement from the SRCCL process at the end. Maybe something to be aware of in further selection of authors.

Final word: I would like to thank the IPCC for having given me the opportunity to contribute to this process, and the WG3 TSU office for its support. It has been a truly enriching experience.

PB

2019-06-10

Final Report of Reviewer Editor

July 6th 2019

By Jhan Carlo ESPINOZA

RE, Ch2 'Land-Climate Interactions', IPCC Special Report on Climate Change, Desertification, Land Degradation, Sustainable Land Management, Food Security, and Greenhouse gas fluxes in Terrestrial Ecosystems (SRCCL)

1.- Overall summary of the chapter team's work in the reviewing process

The chapter team of the Ch2 was composed by expert scientists of land-climate interaction which have large experience in their respective research fields. During the redaction and reviewing processes they have been working in a coordinated and efficient manner. In addition, expert reviewers of worldwide provided suggestions, comments and remarks in each step of the reviewing processes. In consequence from the first version of the Ch2 to the final version, Ch2 was substantially improved.

The FOD received 1981 comments (from 135 reviewers) and SOD received 2917 comments (from 192 reviewers). The percentage of substance comments from total comments for Ch 2 was 50% and 57% in FOD and SOD, respectively. Regarding the FOD, some subsections were under-reviewed, however, the SOD review generated a distribution of comments that is far more homogeneous across all sections and subsections.

Some relevant comment needed substantial attention of the authors such as:

- Text complicated and too long in some subsections.
- Overlaps between chapters, and between sections.
- Overlapping and repetition in some parts of the chapter.
- Over-reliance on model results.
- Unequal inclusion of systems components across the sections: forest management, agricultural practices, biophysical feedbacks.
- Reliance on too few references or on older references in the FOD.
- Issues with assessment language.

All of these issues were well addressed by the authors team, including many others more specific points.

2. Issues required serious attention of Ch2 team at the final stage of the preparation of SRCCL

Since the end of the LAM4 authors worked on a new version of the chapter including suggestions from reviewers on the SOD. The more relevant improvements included restructuration of the chapter in order to avoid repetitions and overlapping. One subsection was removed, and the key messages were moved to other sections. Authors have conducted this process in a very organized manner, which required additional efforts for authors and also for REs in the track of the response to reviewers' comments.

Other improvements were done, providing assessment (in addition to review of literature) and including new scientific references.

Two additional points have received serious attention of the chapter team:

- Including additional information regarding regional particularities of land-climate Interactions.

- Avoid unnecessary technical information in the ES.

I was very pleased to participate in the production of the Ch2 of SRCCL, which in my opinion provide unprecedented assessments of our understanding of land-climate interactions. During this process, it was really a pleasure for me to work with my RE colleagues, Pierre BERNIER and Sergey SEMENOV, and Ch2 authors. Finally, I would like to thank Sarah CONNORS and Andreas FISCHLIN for facilitating the work.



Jhan Carlo ESPINOZA,
IPCC SRCCL Ch2, RE.

FINAL REPORT

by Sergey SEMENOV, RE, Ch2 'Land-Climate Interactions' ,
IPCC Special Report on Climate Change, Desertification, Land
Degradation, Sustainable Land Management, Food Security, and
Greenhouse gas fluxes in Terrestrial Ecosystems (SRCCL)

1. Remarks on the chapter team's work in reviewing and addressing comments at all stages of the preparation of SRCCL

The chapter is of high importance for the report and for the modern climate science in general. It is a successor of the IPCC Special Report on Land Use, Land-use Change, and Forestry (2000). Twenty years passed since that. Reviewing and assessing the new knowledge, which has emerged since that time, is one of the major achievements of SRCCL

The chapter team consists of worldwide known experts in the field from different regions. They have been working in a very efficient manner and in good spirit of cooperation. This allowed to prepare the text of high scientific quality in accordance with the outlines approved by the Panel and general principles for the preparation of the IPCC reports. The main findings and conclusions are supported by relevant literature cited.

From the very beginning of the work, from the FOD, the authors faced some serious challenges reflected in the comments. Amongst them:

- The page limit for the chapter was substantially exceeded;
- The text was rather a review (sometimes even a textbook) than an assessment;
- Policy relevant findings needed more attention;
- Confidence and/or likelihood levels of some statements were missed;
- Some statements were supported by too poor amount of literature;
- Some overlaps between sections of Ch2 (also Ch2 and other chapters) were observed.

The author team has successfully managed with these and many others problems while addressing the reviewers' comments during the editorial period in winter and spring of 2019.

2. Issues required serious attention of Ch2 team at the final stage of the preparation of SRCCL

The final stage of the preparation of Ch2 was not easy. The SOD received substantially more comments than the FOD. Many of them were of substantive nature, some recommended to involve more processes into the consideration. In many cases, the latter was not possible at the final stage of the work or even not

needed at all. The authors have responded to the comments and taken them into account in a very balanced manner.

The team has decided to restructure the chapter in order to avoid overlapping. As a result, one section was omitted, and respective bits of information were moved to other sections, or references to other chapters of the report were made.

I would like to particularly emphasize two points:

- Land impacts on climate and weather through biophysical and GHGs effects;
- differences in land-atmosphere GHG fluxes estimated with different methodologies, in particular, Dynamic Global Vegetation Models (DGVMs) and bookkeeping models.

These issues are extremely important both theoretically and practically. They have been considered by the team in a very fundamental manner with the use of relevant modern literature.

In some cases, problems with uncertainty qualifiers arose. However, needed expert judgments were finally made by the group in the frameworks of existing IPCC approaches (which need some refinement, I think).

Ch2 team has made substantial efforts towards the proper reflection of the major statements in the ES and SPM. As a result, the main findings and conclusions of Ch2 are presented in the SPM accurately and in objective and balanced manner.

It was really pleasure for me to work with Ch2 authors, notably CLAs Elena SHEVLIAKOVA and Gensuo JIA, my RE colleagues Pierre Bernier and Jhan Carlo Espinoza, and the IPCC TSUs members facilitating the work.



Sergey SEMENOV,
IPCC SRCCL Ch2 RE

Final Report on IPCC SRCCL, chapter 3 „Desertification“

Dr: Mariam Akhtar-Schuster, Review Editor Ch. 3 „Desertification“ (5th July 2019, Berlin)

Responses to queries: Do any significant differences of opinion on any scientific issues that arose during the drafting and review processes exist? How were these issues reconciled in the final text?

The following scientific issues became visible during the review processes of the FOD and SOD. Although these have been carefully analysed and addressed by the authors of ch 3, they may require further attention in the final stages of the process:

• **Definitions used to define desertification and land degradation:**

Comments provided during the review process asked for clarification on how ‘desertification’ is defined. Clarification on this issue was crucial as it was the basis for understanding (a) the scope of ch 3, particularly providing a distinction between desertified areas and deserts, (b) the reason for having a chapter on ‘desertification’ (ch 3) and a chapter on ‘land degradation’ (ch 4), and (c) the relationship between ch 3 and the other chapters of the report, particularly chapter 4.

- Several comments received during the FOD and SOD review processes addressed aspects regarding the definition of desertification. These were analysed by the authors in a dispassionate scientific atmosphere (I attended the 3. and 4. Lead Author Meetings (Dublin & Cali)).
- The analyses undertaken by the authors showed that these comments generally did not contradict the UNCCD definition, which had been used. Rather, these comments provided aspects to further underpin the definition used. The authors considered the comments received to further frame their discussions. This effort is reflected particularly in section 3.2.1.
- The Executive Summary builds on these analyses by starting with the sentence that desertification is confined to arid, semi-arid and dry sub-humid area, collectively known as drylands. This provides a clear distinction to ch 4 that covers land degradation in other climate systems.
- The introductory chapter also clarifies that deserts (“valuable ecosystems”) are located in drylands. They are however not considered prone to desertification. Thus, desertification cannot be equated to desert expansion. With this scientific distinction, the authors have also carved out the vital economic, cultural and ecological role which favourable isolated areas play for production in hyper-arid/desert areas (e.g. oasis with their reliable water supply) as well as their exposure/sensitivity to unsustainable land use and climate change. These aspects could be useful for aspirations towards sustainable development in hyper-arid regions.
- It may be important to underscore that the definitions used for defining desertification and land degradation (ch 4) are complementary. Both definitions show that “the difference between desertification and land degradation is not process-based but geographic.” Therefore, both chapters strongly relate to each other. This may also clarify, why certain aspects are logically placed in either of the chapters (also to avoid lengthy duplications). For instance introducing the LDN conceptual framework in ch 4 is logical because the concept applies to degradation under all climate regions, including drylands. It is therefore sufficient that it is described once in the special report. On the other hand, the authors of ch 3 and 4 jointly developed a cross-chapter box (see ch 3) on policy responses to drought, a climate phenomenon which is not only restricted to drylands (this sharing of knowledge shows the inherent relationship between both chapters).

Possibly remaining issues: Regarding the numerous comments on definitions and boundary issues, it may be useful to develop a crisp scientific message on the relationship between desertification, land degradation and deserts, possibly also including the issue of droughts. At the same time the intrinsic thematic relationship between chapters 3 and 4 may need to be kept under consideration: Certain aspects may be discussed in detail once under either of the chapters.

• **“The AI [Aridity Index] is not an accurate proxy for delineating drylands in an increasing CO2 environment”**

This is possibly an important statement emerging from ch 3 for many outside the strictly ‘climate community’. Discussions on the implications of rising CO2 levels to estimate dryland areas may need a

brief and clear description because considering the AI has been decisive in dryland /desertification discussions during the past decades in various land-related scientific communities.

- **The issue of quantitative data**

There were many questions/requests during the FOD and SOD review processes to include more quantifiable information. The authors have included evidence-based quantifications whenever available. The confidence levels in ch 3 clearly reflect the available evidence base. This makes the message emerging from this ch solid.

(Further thoughts: Different disciplines have different methods. This may require considerations in future on how to manage more qualitative messages in IPCC assessment processes. Qualitative statements/messages that are based on a solid narrative could provide a way forward.)

- **“Limits to adaption”**

Comments regarding “Limits to adaptation” have been addressed. There is however insufficient knowledge available on this matter. It may need to be considered that the term “limits to adaptation” is being used as a scientific concept by the authors, and the comments received on the FOD and SOD on this subject are also of scientific nature. This issue may just as well have been termed “limits to adaptive capacity” (see the use of the term “adaptive capacity in e.g. SOD Review comment 12441, 12443).

- **Addressing regional imbalance**

The definitions used in ch 3 (see introduction) provide clarification for the selection of the case studies presented in this ch. Based on numerous comments emerging from the FOD and SOD review processes, the authors have improved the regional balance, particularly also through diversifying the set of case studies included in the chapter. These improvements were based on extensive scientific discussions, which took place at the 3rd and especially 4th Lead Author Meeting and extensive literature research in the immediate aftermath of these meetings. The inclusion of case-studies from hyper-arid regions (example Oasis) has closed a crucial regional gap in ch 3. This chapter now also addresses West Asia and the Arabian Peninsula (see also my comment above on the economic, cultural and ecological relevance of favourable isolated areas in hyper-arid environments).

- **Sustainable land management (SLM):**

SLM is discussed/used widely in the land-related scientific communities. SLM was considered to be an overarching theme during the scoping process for this special report, an explanation for why it was not placed in one chapter as is the case for ‘desertification’ (ch 3), land degradation (ch 4) and food security (ch 5). In the SOD review process, the number of references to SLM had visibly increased.

In order to ensure a solid understanding of the SLM approach in the further process, it may be useful to provide brief comprehensive context-specific explanations regarding SLM. This could be done jointly by the authors of ch 3 and ch 4, who both discuss SLM.

Further, SLM can help to avoid desertification, and support measures to reduce desertification. SLM has the potential to deliver on multiple SDGs. It can support aspirations for sustainable development to ensure human wellbeing – worldwide, particularly under ongoing or anticipated climate change effects. It may therefore be useful to reflect on SLM more strongly as a major policy-relevant “umbrella” option for land-based climate change adaptation and mitigation in the Summary for Policymakers; ch 3 findings can support this.

SLM could also support the development of further science interactions between ‘land’ and ‘climate’ experts.

General comment: It was a pleasure for me to be involved in this inspiring work and observe with how much commitment and dedication the authors worked on ch 3, thereby analysing the numerous comments emerging from the FOD and SOD review processes to further develop this chapter.

M. Akhtar-Schuster

Dr. Mariam Akhtar-Schuster (Berlin, 5th July 2019)

Review editor report on Chapter 3 of the IPCC special report on climate change and land (IPCC Special Report on climate change, desertification, land degradation, sustainable land management, food security and greenhouse gas fluxes in terrestrial ecosystems)

Submitted by: Fatima Driouech

- Casablanca, 07/07/2019

Hereafter I summarize the main review activities I undertaken by myself and the other two review editors of chapter 3 and my conclusions issued from this review.

The review activities included a careful reading of the first and the second order draft of the chapter, the analysis of the comments received for each of the versions and the evaluation of the author team responses and associated redrafting in the chapter. Interactions and exchange with the chapter team during the 3rd and the 4th author meetings allowed working more closely with the authors in treating the comments, especially the substance and uncertainty languages categories starting with the more critical ones. A special interest was given to how comments/suggestions were addressed, which were rejected by the authors and how the rationale provided for the rejection was solid and comprehensive. We also interacted with the chapter team to ensure that the responses are drafted on an appropriate way.

The comments received and their careful consideration by the chapter team allowed good improvements of the chapter including clear differentiation between the purpose of chapter 3 and chapter 4 (thanks also to the fruitful cooperation between these two chapters), more regional balance in the assessment, the highlighting of knowledge gaps and the assessment of relevant work carried out by other UN organisations. Thus the chapter do not duplicate efforts but can be considered complementary to recent other reports on the subject of land degradation and restoration. Taking into account different review comments, the authors enriched the chapter by the inclusion of additional aspects/topics as such as groundwater and irrigation, more emphasis on gender and poverty, dust and sand storms, climate services and ecosystem services. Future projections based on scenarios have been included for population exposure and discussed for drivers of desertification. Future changes of desertification depending on scenarios could not be included consistently due to lack of publications addressing the topic (limited to very few areas). Limits to adaptation and adaptive capacities have been mainly discussed from the scientific perspective. The Lack of scientific evidence that emerged in some cases has not affected the quality of the messages provided by the chapter.

The process review resulted in an appropriate chapter revision and the treatment of all comments by the authors who took into account a vast majority of them and provided a rationale underlying their decisions of acceptance or rejection.

In my judgment, I feel that the review comments have been sufficiently addressed in terms of both the responses and redrafting.

A handwritten signature in blue ink, consisting of several vertical and curved strokes, appearing to be the initials 'FD'.

Fatima Driouech

Review Editor's report: IPCC special report on climate change and land

Chapter 3: Desertification

Mahesh Sankaran, National Centre for Biological Sciences, Bengaluru, India

Chapter 3 deals with the issue of desertification under climate change, and the reciprocal nature of linkages between the two. In total, there were over 2300 comments from well over a 100 reviewers on the first and second order drafts of Chapter 3. The bulk of these comments (nearly 60%) addressed issues relating to the substance of the chapter, while the remaining was largely editorial or pertaining to the uncertainty language used. There were also several positive comments on the chapter content. Overall, the authors have done a fairly thorough job of dealing with these comments. As a result, the chapter has been considerably strengthened, and is now more focused, crisp and internally consistent. In addition, the chapter is now also better aligned with other similar assessments such as the IPBES Land Degradation and Restoration Assessment.

Specifically, the Executive Summary is now considerably improved. The language has been simplified and it is a lot easier to read. Uncertainties and confidence levels associated with estimates have been provided where relevant, as well as future projections where appropriate. The ES also now discusses gender issues, which was a concern raised earlier. In terms of the main body of the chapter, it now contains more detailed discussions on limits to adaptation, climate services and poverty. Policy prescriptive language has been removed, and confidence language improved throughout. There is also better cross-referencing between chapter 3 and other chapters of the report.

The authors have also made a commendable effort to minimize confusion associated with the definition of terms. In particular, desertification – as different from degradation in general, and from aridity – is now clarified better. Also, the authors have done a good job in clarifying issues associated with the use of the Aridity Index for future projections of desertification – comments relating to these aspects in the previous draft have been addressed satisfactorily. Regional coverage is also much improved (within the limits of the word constraints in the document), which was another major area of concern based on the comments received.

There were also several comments relating to the issue of baselines. The revised draft now includes text at the start which acknowledges issues with, and the importance of, specifying baselines. The authors clarify at the start that different studies use different baselines (typically the start of the assessment period for quantifying change), and acknowledge that providing estimates of the extent of desertification based a common baseline is not possible. In addition, the authors also acknowledge the extensive discussion on 'baselines' that is available in the IPBES LDRA. Overall, this chapter builds on, and in many ways complements the IPBES LDRA –

in particular, the interaction between climate change and degradation in drylands (desertification) is explored in greater depth here in accordance with the mandate of this report.

The chapter now contains a more nuanced discussion of desertification and conflict, as well as a more detailed consideration of issues relating to groundwater and rainwater harvesting. Potential solutions to address problems of desertification (e.g. SLM) have been highlighted where relevant. Finally, several complex sentences in earlier drafts have now been simplified making the current version much clearer and easier to read.

Overall, the authors have done a commendable job in addressing and responding to comments. Many comments that were not addressed (i.e. response noted as rejected) were comments that were either incomplete, referred to the wrong line/ page numbers, were policy prescriptive or, in some instances, not entirely relevant or within the scope of the chapter. The authors have largely done a good job of responding to such comments and have provided valid reasons for not accommodating suggestions.

Some potentially outstanding issues include:

- Afforestation is talked about as a potential solution in places (towards the end of the document/ case studies) although it can also be a cause for degradation (as the authors have noted earlier in the document). This can be a potential source of confusion for readers and policy makers. Although a lot of this text has been added in response to comments received on earlier drafts (positive effects of tree planting), these could have been accompanied with caveats pointing out the potential drawbacks of such initiatives where relevant.
- In some instances, the author responses to comments (in the spreadsheet as opposed to the main text) could have been more detailed. For example, in a few instances the author responses just say 'noted', and it is unclear exactly what action was taken to address the comments. More elaborate and clearer responses would have helped in these instances. Also, responses to some comments could have been more diplomatic.

Mahesh Sankaran

A handwritten signature in blue ink, appearing to read 'Mahesh Sankaran', with a horizontal line underneath.

July 5, 2019

FINAL REVIEW EDITOR REPORT OF CHAPTER 4 OF IPCC SPECIAL REPORT ON CLIMATE CHANGE AND LAND (SRCCL)

1. Main statistics

The chapter has undergone two rounds of review, a first one by experts only and a second one by experts and governments. The following table summarizes the number of reviewers that submitted comments, the total number of comments received and their nature, based on the categorization provided by the reviewers themselves.

Table 1. Main statistics about the reviewers and comments received by Chapter 4 of the SRCCL.

| Draft | Number of Reviewers | Total comments | Type of comment | | | |
|--------------|---------------------|----------------|-----------------|-----------|------------|----------------------|
| | | | Substantive | Editorial | References | Uncertainty language |
| FOD | 84 | 1.115 | 722 | 262 | 87 | 44 |
| SOD | 128 | 1.745 | 1.178 | 361 | 151 | 55 |
| Total | 212 | 2.860 | 1.900 | 623 | 238 | 99 |

In both reviews, a large number of comments were substantive or provided additional references to the authors. The comments affected all parts of the chapter, which means that text was not only rather extensively commented on, but that the review was rather complete.

2. Topics of concern during the review

Following are some of the topics that appeared more frequently in the comments by the reviewers. They are listed in alphabetic order, without assigning to any of them more importance than to the any other:

- Clarification of the divisions of land being used
- Clarification of the naturalist and social scientist views about land degradation
- Clarification of the role of biophysical and socioeconomic factors as drivers of land degradation
- Clarifications of a number of concepts being used
- Heterogeneity among sections in the style of writing and depth
- Incorrect statements not supported by the literature
- Key messages needed to be sharpened
- Lack of examples or good examples, particularly in the case studies
- Lack of support or limited support for a number of paragraphs, at times with outdated references
- Lack of temporal and scenarios frameworks
- Lack or insufficient geographic details
- Limited cross- referencing with other chapters of SRCCL and other IPCC reports
- Limited use of the uncertainty language
- Much of the text is written in review mode rather than as an assessment
- Overlapping text with other chapters
- Poor or insufficient artwork of figures or incompleteness of these and tables
- Review was often incomplete and, at times, biased

- Unclear differentiation between Chapter 3 and Chapter 4
- Use of policy prescriptive language

3. Main contentious issues

- The definition of land degradation received many comments, at times contradictory between reviewers
- The separation between non-forest land and the forest-land

4. Analysis of the responses to the comments

The author team worked during the FGD preparation period, providing appropriate answers to all comments. The text was changed taking into consideration the comments made in a majority of cases when substantive comments were made, only a few comments were rejected based on solid grounds. During LAM4 the contentious issues were discussed at length and in coordination with other chapters, to clearly draw the lines of the concepts being used and to reduce overlapping among chapters. Finally, the team focused in rewriting the chapter executive summary, improving or incorporating the use of uncertainty language and assuring the traceability of the conclusions with the chapter text.

5. Conclusion

In my opinion, all substantive and contentious issues raised by the reviewers received appropriate consideration, and resulted in the corresponding clarifications and amendments in the text when needed. The review was successfully completed.

Toledo, July 12nd, 2019

A handwritten signature in black ink, reading "José M. Moreno", enclosed within a hand-drawn oval shape.

José M. Moreno

Review Editor of Chapter 4 of the SRCCL

FINAL REVIEW EDITOR REPORT OF CHAPTER 4 OF IPCC SPECIAL REPORT ON CLIMATE CHANGE AND LAND (SRCCL)

1. Main statistics

The chapter has undergone two rounds of review, a first one by experts only and a second one by experts and governments. The following table summarizes the number of reviewers that submitted comments, the total number of comments received and their nature, based on the categorization provided by the reviewers themselves.

Table 1. Main statistics about the reviewers and comments received by Chapter 4 of the SRCCL.

| Draft | Number of Reviewers | Total comments | Type of comment | | | |
|--------------|---------------------|----------------|-----------------|-----------|------------|----------------------|
| | | | Substantive | Editorial | References | Uncertainty language |
| FOD | 84 | 1.115 | 722 | 262 | 87 | 44 |
| SOD | 128 | 1.745 | 1.178 | 361 | 151 | 55 |
| Total | 212 | 2.860 | 1.900 | 623 | 238 | 99 |

In both reviews, a large number of comments were substantive or provided additional references to the authors. The comments affected all parts of the chapter, which means that text was not only rather extensively commented on, but that the review was rather complete.

2. Topics of concern during the review

Following are some of the topics that appeared more frequently in the comments by the reviewers. They are listed in alphabetic order, without assigning to any of them more importance than to the any other:

- Clarification of the divisions of land being used
- Clarification of the naturalist and social scientist views about land degradation
- Clarification of the role of biophysical and socioeconomic factors as drivers of land degradation
- Clarifications of a number of concepts being used
- Heterogeneity among sections in the style of writing and depth
- Incorrect statements not supported by the literature
- Key messages needed to be sharpened
- Lack of examples or good examples, particularly in the case studies
- Lack of support or limited support for a number of paragraphs, at times with outdated references
- Lack of temporal and scenarios frameworks
- Lack or insufficient geographic details
- Limited cross- referencing with other chapters of SRCCL and other IPCC reports
- Limited use of the uncertainty language
- Much of the text is written in review mode rather than as an assessment
- Overlapping text with other chapters
- Poor or insufficient artwork of figures or incompleteness of these and tables
- Review was often incomplete and, at times, biased

- Unclear differentiation between Chapter 3 and Chapter 4
- Use of policy prescriptive language

3. Main contentious issues

- The definition of land degradation received many comments, at times contradictory between reviewers
- The separation between non-forest land and the forest-land

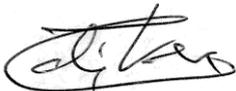
4. Analysis of the responses to the comments

The author team worked during the FGD preparation period, providing appropriate answers to all comments. The text was changed taking into consideration the comments made in a majority of cases when substantive comments were made, only a few comments were rejected based on solid grounds. During LAM4 the contentious issues were discussed at length and in coordination with other chapters, to clearly draw the lines of the concepts being used and to reduce overlapping among chapters. Finally, the team focused in rewriting the chapter executive summary, improving or incorporating the use of uncertainty language and assuring the traceability of the conclusions with the chapter text.

5. Conclusion

In my opinion, all substantive and contentious issues raised by the reviewers received appropriate consideration, and resulted in the corresponding clarifications and amendments in the text when needed. The review was successfully completed.

Buenos Aires, July 3rd, 2019



Carolina Vera

Review Editor of Chapter 4 of the SRCCL



DEPARTMENT OF AGRONOMY
Faculty of Crop Production Sciences
The University of Agriculture Peshawar
Khyber Pakhtunkhwa
Pakistan-25130



FINAL REPORT OF CHAPTER-5 ON FOOD SECURITY

I checked the chapter very carefully. The chapter is now in the most improved form as compared to the one at Dublin and CIAT. All the differences of opinion arose during the drafting and review process are very carefully addressed and incorporated by the authors. I am highly impressed from the combined chapter team's work while reviewing and addressing all comments and suggestions. Moreover, the authors were also most positive to REs comments during the whole process. The authors have replied to all comments of the reviewers in a very positive way. I congratulate the whole team of chapter 5 especially the lead authors **Dr. Cynthia Rosenzweig** and **Dr. Cheikh Mbow** for their positive responses and smiles. By the way, I appreciate the reviewers who actually helped the authors to bring this chapter to the final and useful shape.

I am once gain thankful to IPCC for nominating me as the RE for food security (a major issue of my country, Pakistan). I learned a lot during the meetings in Dublin and CIAT and hope to learn more with cooperation with IPPC, thanks.

Dr. Amanullah

Review Editor Chapter-5 (Food Security)
July 04, 2019

Review Editor Report
Final Draft of IPCC SRLCC
Chapter 5 - Food Security
RE: Prof. Nouredine Benkeblia

After successive editing and corrections, the chapter was significantly improved and the content thoroughly polished when compared to the previous drafts. The final draft looks clear and for sure it will be very useful for all the communities including scientific, stakeholders and policy makers all together.

After hundreds of comments that have been raised by the reviewers, the contributing authors and the lead authors and under the excellent coordination of the Coordinating authors, have addressed most of these comments, while few have been reconsidered and the text was polished. However, few points have not been adequately addressed due to the discrepancies of the reviewers' comments and lack of information and data to clarify these comments. Therefore, I noticed that the authors and contributors addressed these comments in a succinct and precise manner when possible.

Therefore, I can conclude that:

- The responses to the reviewers; comments are concise and more than to be satisfactory.
- The text is clear and well supported by the scientific evidence.
- Tables and Figures are appropriate and explanatory and supporting well the content of the text.

- The data reported are relevant and significant.
- The quality of the reported data seems to be technically sound, appropriate techniques have been cited and described
- The reported data have been well and critically analysed and interpreted.
- The data have been clearly presented and sufficient detail given to be clear and understood by the readers.
- The level of support for the conclusions is sufficient and strong evidence is provided for the authors' claims.
- The reported data are very important and useful to the field.
- The document is well streamlined.

Overall, the final draft might be considered as final version, and might be approved under its present form.



N. Benkeblia

RE: Prof. N. Benkeblia

Review Editor Final Report
Chapter 5

Andy Challinor 1/7/2019

The hard work of the authors clearly improved the report from its very sound starting point. It is clearly something to be proud of. In general, responses to comments are concise, relevant and firm where necessary.

At the FOD stage, the authors focussed the chapter on climate change and food security, rather than simply food security as a topic. The authors also provided a greater focus on regional aspects, as recommended in several comments. Awareness of the important role of uncertainty language also came in at this stage, in response to comments. The authors accepted a very high percentage of the FOD comments from reviewers, whilst being appropriately wary of responding to comments that sought to add citations.

At the SOD stage, the chapter was again improved and polished from the previous version. The framing figure was again been improved; no doubt it will prove useful elsewhere also. The wording of the ES has matured nicely into a nuanced but firm set of evidence-based conclusions, perhaps partly in response to rejections of stronger wording in the SOD.

Especially useful review comments at the SOD led to:

- Greater emphasis on the role of education, via specific text in the capacity building section.
- The adaptation section (5.3) was vastly improved, with greater background, framing and lead-in to the details that were present in the SOD. Relevant detail has also been added, for example on risk management.
- Overall, the document has been further streamlined, whilst still managing to respond to reviewers' comments. For example, a comment (21491) was made about the use of novel technology to reduce livestock emissions, stating that the two lines devoted to it were not enough. A signpost to another section and the addition of supplementary information (SM5.5) elegantly dealt with this issue.

To the IPCC Working Group III Co-Chairs

Final Report from Review Editors for Chapter 6 of IPCC Special Report on Climate Change and Land (IPCC SRCCL: Chapter 6: Inter-linkages between Desertification, Land 2 Degradation, Food Security and GHG fluxes: 3 synergies, trade-offs and Integrated Response 4 Options.

After the Second Order Draft (SOD), when over 1430 comments were received, and two of the Review Editors was unable to attend the Forth Lead Author meeting, it became clear that the work load had become unmanageable and more Review Editors were requested.

The TSU reacted positively. I was approached by TSU and Co-Chairs of WGIII to be additional Review Editor, as I have already been following the chapter during LAM3, therefore was in a good position to take over this task.

CLAs and LAs spend significant effort and ample time at the LAM4 addressing the ES and key messages/storyline arising from the chapter, working across chapters and cross chapter boxes and contributing to outlining the SPM. The Author Team worked very well as a team and the chapter took good shape.

Overall comments:

- The CLA's and LA's have responded appropriately, and where necessary in detail to the First Order Draft (FOD), which received 1128 reviewer's comments, while for the Second order Draft SOD received ~1430 reviewers' comments. Many of these are generally constructive, pointing out areas for further improvement and clarification.
- Based on the SOD and the review comments received, the Review Editors have highlighted a number of issues that seemed to be critical for a successful chapter, or were particularly pushed by the reviewers.
- The reviewers have commended the chapter for the overall high quality, major improvement since FOD, despite the extremely ambitious time frames and very high number of comments.
- The reviewers noted that there were number of repetition with other chapters, and some cross-referencing would be useful. And lack of information in the executive summary about adaptation options, and statements reflected were rather general.
- There were number suggestions to paraphrase the language where layman's could easily be read in the SPM.
- The reviewers noted that the chapter would overall benefit from tightening up the content and cut some redundant material already mentioned
- The reviewers welcomed the crossover approach used in Chapter 6. and suggested that the analysis of the available scientific literature be refined, since several additional studies could be used to better assess the effects of some land-based options on other sustainable development objectives, for example on the links between REDD+ and

food security, or the adaptation benefits associated with material substitution. These studies are indicated in various comments made throughout Chapter 6.

- The reviewers noted that the need to complement the assessment presented in chapter, in particular in section 6.4., with one more response option: The option of "inaction" beyond actions in place today. This would put the information and figures in context and convey the message that although some of these response options do come at a certain disadvantage doing nothing might not be an option.
- Some reviewers pointed out that the Ch. 2-5 would bring up a number of factors and feedbacks that are involved in climate forcings, including albedo, water retention, carbon retention and methane decomposers in upland soils. For instance, these are summarized in section 2.6.2.1. Thus, they would also know that these cycles could be managed for the sake of mitigation. However, in the holistic assessment of various response options in ch.6. (which is generally very much appreciated), perspectives on mitigation falls down to GHG emissions only.
- The reviewers pointed out to consider including biogeophysical factors that are relevant for mitigation, as well as difficulties over metrics to represent such factors.
- The reviewers also recommended to consider the consistency between the different chapters related to strategies for livestock management. Mostly, the idea for "improved livestock management" in ch. 6 (see for instance 6.3.2) is to reach higher yield per unit of input/emissions. Thus, there is an idea for "output-optimization". Similarly, in discussions over diets in ch. 5 and 6, assessment is also exclusively output-oriented, focusing on emissions per unit of output.
- They also pointed out on the other hand, the livestock sector is naturally also at the center for proposals "improved grazing land management" and "avoided conversion of grassland to cropland". Further, in chapter 5 (See 5.5.1.4,5.6.3) we are presented with "integrated responses to crop and livestock". In these proposals, some of the idea is that livestock can utilize resources that are otherwise wasted.
- They also recommend to consider a cross section box on livestock assessing merits of an output-optimized approach (i.e. emissions per unit output) compared to an input-optimized approach (i.e. including grazing land management, grazing strategies that are helpful for soil carbon, use of marginal resources, use of water and grasses from regions where such resources are plentiful etc.).
- Reviewers also pointed out the Inter-linkages between desertification, land degradation, food security and GHG fluxes: synergies, trade-offs and integrated response options.

The Chapter has improved considerably taking into all substantive and editorial comments from reviewers and the chapter is in a pretty good shape.

Signature:



Full Name: Amjad Abdulla

Date: 20 July 2019

Report by Professor Ian Noble, Review Editor for IPCC SCRRL Special Report
Chpt 6

“Interlinkages between Desertification, Land Degradation, Food Security and
GHG fluxes: synergies, trade-offs and Integrated Response Options”.

I attended the 3rd and 4th Lead Author Meetings in my role as Review Editor (RE) of Chpt 6 of the above IPCC Special Report. The chapter has three REs, Professor Yoshi Yamagata, who attended the 3rd LA meeting, Dr Taha Zatari, who was not able to attend either meeting, and me. All REs reviewed the comments and the authors’ responses for both meetings.

This chapter deals with the “Interlinkages between Desertification, Land Degradation, Food Security and GHG fluxes: synergies, trade-offs and Integrated Response Options”. As such it brings together the interactions between 40+ key issues discussed in the four preceding chapters.

The work was well led by the two CLAs with the participation of the entire writing team and the Chapter Scientists. At the meetings, both Prof Yamagata and I were welcomed into the drafting team and were able to provide advice while also confirming that major comments from reviewers were being taken into account.

There were 1128 comments on the Second Order Draft and 1430 for the Third. Most comments were on minor editorial points, points seeking greater clarity and suggestions for additional references. However, a common theme through both meetings was that the chapter had a very complex structure and was difficult to follow. The authors were already very much aware of its complexity and spent significant time at both meetings, and between meetings, seeking to simplify the structure to improve its readability and relevance to policy formulation.

My assessment is that the authors dealt with the comments effectively. They made extensive changes to the chapter in terms of clarifying wording, better referencing and in restructuring many sections. They also worked with other chapters to improve cross referencing and to reduce duplication.

There were no major contentious issues although on a few topics (e.g. biochar) differences in scientific assessment showed through, but the authors dealt objectively with those comments in their responses. The authors also clarified and justified their use of the relatively unfamiliar concept of “anthromes” as the most effective way of dealing with the global patterns of human-altered ecosystems.

In the Third Order Draft some reviewers still differed in their view about the coverage and structure of the chapter and sometimes it’s relationship to other chapters. A few provided multiple comments on this point sometimes with suggested alternatives. Many of these suggestions would have meant a major restructuring and redrafting of this, and sometimes other, chapters. At this stage the chapter has been through multiple reviews and by hundreds of reviewers. There is no right or wrong structure for a chapter with such cross-

cutting emphasis as Chpt 6, and the comments did not reflect on the accuracy or the main messages of the report. In my opinion the authors, selected for their expertise in this area, made significant structural changes to simplify and clarify the chapter but have had to reject the more unrealisable alternatives.

I commend the authors and Chapter Scientists for their efforts to bring together a difficult chapter and their diligence in taking account of reviewers' comments.

A handwritten signature in blue ink that reads "Ian Noble". The letters are cursive and fluid.

25 July 2019

Professor Ian Noble, Review Editor

Report on Final Review Comments Chapter 6

I have reviewed the responses to the reviewers' comments.

My view is that the chapter final draft had responded to the comments adequately. The final draft has improved greatly by revising the sections and the explanation become clear. Reviewers comments are different in their views on the structure of chapters, but authors responded adequately by clarifying the descriptions of each chapters.

In this chapter, there are some policy relevant important messages regarding the whole report such as the SDGs implications of the sustainability of BECCS. There are some policy relevant sensitive issues regarding the differences among scientific literatures.

The authors successfully created the final assessment using useful tables and figures that integrate various assessment results. Especially, I view that it is important that implications of large-scale biomass implementation were included in the final report with enough number of literature reviews.

Yoshiki Yamagata

A handwritten signature in black ink, appearing to read 'Yamagata', written in a cursive style.

Review Report on Chapter 6 of SRCCL Interlinkages between Desertification, Land Degradation, Food Security and GHG fluxes: synergies, trade-offs and Integrated Response Options

Taha M. Zatari

I have reviewed final draft of the sixth chapter of the IPCC Special Report on Climate Change and Land (SRCCL).

The chapter illustrates interlinkages between sustainable land management options to deliver climate change mitigation and adaptation, to prevent desertification and land degradation, to enhance food security and to assess reported impacts on ecosystems services and their contributions to the UN SDGs.

The authors, during the process of development of this chapter, have taken into consideration and responded to all of the editorial, substance, references and language uncertainty issues raised by reviewers and addressed them positively and objectively. The following points have been noted.

- The authors have responded to all of the 101 reviewers' 1430 comments which included 894 substance, 429 editorial, 68 references and 39 uncertainty language comments.
- The authors have accepted and positively addressed a large number of those comments. Nearly all of the editorial and reference comments have been addressed by the authors.
- This final draft of the chapter is a significant improvement over its previous drafts.
- The executive summary has been modified to be more explicit and inclusive. More examples and quantification statements have been added.
- The issues of repetitions within and with other chapters raised by reviewers have been addressed.
- The authors have revised chapter 6.4 to be quantitative in approach based on the recommendations of the reviewers.
- The text in reference to the tables and figures in the chapter has been expanded to clarify the theme.
- Some of the subheadings in the chapter have been revised to shorter and easy to understand subheadings on the reviewers' suggestions.
- Reviewers' suggestions regarding uncertainty statements have been revisited and revised where strong conclusion was drawn based on weak or limited evidence. All such statements have been revisited and revised.

- A huge number of rewording for whole chapter including executive summary has been accepted and adopted by the authors.
- Overlapping issues have been addressed by cross referencing.
- Figure 6.1 has been redrawn and new caption added to address carbon fluxes and non-CO₂ GHG factors influencing climate.
- Risk management has been added in chapter.

The literature survey has been further expanded and strengthened to cover this vast subject and to give credibility to the chapter and its inferences. Some issues have been dealt in more details than others.

The authors have succeeded in their effort of structuring a chapter of complex issues into a simpler and understandable form. The figures have been made simple and reflective of the contents.

The contents of chapter 6 cover all of the topics proposed as outline of the chapter adopted by the IPCC at its 45th Session. However, the size of the chapter has exceeded the allocated limit of ~40 pages. Although, I can understand that these vast subjects outlined above can't be covered in just ~40 pages

The chapter has included five case studies (Box 6.1 A-E) located in different world regions as stated in the outline, but have not included other region(s). Some parties have asked for regional balance by covering other regions in the case studies.

The overall chapter in its current form is satisfactory and addresses all of the relevant issues.

Taha M. Zatari
Review Editor, Chapter 6, SRCCL

Final Report - IPCC SRCCL - Chapter 7

Review Editor: Regina R. Rodrigues

Several reviewers noted that the section on decision making was general in kind and should be concrete with examples as the concern to land-based decisions and that a case study would prove useful. However, this was challenging as other reviewers had asked to be more generic and to avoid using specific examples. In the end, the Coordinating Lead Authors and Lead Authors managed to balance the two approaches by including concisely in the text generic findings relevant to many countries, leaving some more specific examples in boxes, as case studies.

There was also criticism by some reviewers of the definition and discussion on risk during the first round of review, related to the First Order Draft. But once again, the Coordinating Lead Authors and Lead Authors, with the help of the Technical Supporting Unit, handled this very well. Extensively discussed during the 3rd Lead Author Meeting in Dublin provided a standard definition of risk that was consistent with previous IPCC reports and used in other SRCCL chapters.

Vast improvement was also made in relation to the Figures and Executive Summary of Chapter 7 from the First and Second Order Draft to the Third Order Draft, in response to specific requests from many reviewers.

Therefore, in my view, the work of the Coordinating Lead Authors and Lead Authors on responding to the review comments was excellent.



Dr Regina Rodrigues Rodrigues
Federal University of Santa Catarina
Department of Oceanography
Florianópolis, SC, Brazil

Chapter 7 Comments: B. L. Turner II

Overall, C7 captures the state-of-the-art/science as it applies to its assigned topic. It lays out the main issues and problems of climate change in the land sector, foremost those identified through international programs assessing the impacts of and possible responses. It is noteworthy that C7 treats the policy dimensions (7.5) fairly, recognizing that research community at large is not sufficiently advanced to prescribe the level of specificity about evidence and agreement that can be applied in other sections of IPCC considerations. (Note that my core expertise resides in those sections of C7 section 7.5, whereas that for 7.5 I assign myself modest expertise.)

A few issues to consider.

Throughout C7 there are sentences/phrases that are correct, assuming appropriate interpretation of them. These could easily be rephrased to add clarity. Examples include:

- Page 4, line 10: “components” apparently refer to the list of items preceding the sentence but those items were not identified as components.
- Page 4: Line 22-23 appears to have a missing word: “inadequately shielded from” should be “shielded them from”
- Page 13: Line 23: all people are dependent on agricultural production; I think the intent is to say that these people are dependent on local or subsistence production.
- Throughout are various statements that “policies” can deal with x or y, but as stated imply that “any” policy can do so, as opposed to appropriate policy.

7.3.2.2 No mention is made of the positives that may come from in *international* food linkages, especially when they serve as safety nets as in famines. Note that on page 60, line 23, the positives of *global* safety nets would appear to contradict the absence of any positives from international food linkages.

7.3.2.7 There are very recent reports that permafrost thawing is well ahead of projected estimates. Perhaps this evidence is too recent to include.

7.5.4.2 carbon pricing and 7.5.6.2. PES are well treated but do not identify as an issue the inappropriate cost for carbon or other services that follow from adding non-carbon and non-services to policy, such as REDD+.

IPCC Special Report on Climate Change, Desertification, Land Degradation, Sustainable Land Management, Food Security, and Greenhouse gas fluxes in Terrestrial Ecosystems

**Summary for Policymakers
Approved Draft**

Headline Statements

Subject to copy edit and layout

A. People, land and climate in a warming world

A 1. Land provides the principal basis for human livelihoods and well-being including the supply of food, freshwater and multiple other ecosystem services, as well as biodiversity. Human use directly affects more than 70% (*likely 69-76%*) of the global, ice-free land surface (*high confidence*). Land also plays an important role in the climate system. {1.1, 1.2, 2.3, 2.4, Figure SPM.1}

A 2. Since the pre-industrial period, the land surface air temperature has risen nearly twice as much as the global average temperature (*high confidence*). Climate change, including increases in frequency and intensity of extremes, has adversely impacted food security and terrestrial ecosystems as well as contributed to desertification and land degradation in many regions (*high confidence*). {2.2, 3.2, 4.2, 4.3, 4.4, 5.1, 5.2, Executive Summary Chapter 7, 7.2}

A 3. Agriculture, Forestry and Other Land Use (AFOLU) activities accounted for around 13% of CO₂, 44% of methane (CH₄), and 82% of nitrous oxide (N₂O) emissions from human activities globally during 2007-2016, representing 23% (12.0 +/- 3.0 GtCO₂e yr⁻¹) of total net anthropogenic emissions of GHGs¹ (*medium confidence*). The natural response of land to human-induced environmental change caused a net sink of around 11.2 GtCO₂ yr⁻¹ during 2007-2016 (equivalent to 29% of total CO₂ emissions) (*medium confidence*); the persistence of the sink is uncertain due to climate change (*high confidence*). If emissions associated with pre- and post-production activities in the global food system² are included, the emissions are estimated to be 21-37% of total net anthropogenic GHG emissions (*medium confidence*). {2.3, Table 2.2, 5.4}.

A 4. Changes in land conditions³, either from land-use or climate change, affect global and regional climate (*high confidence*). At the regional scale, changing land conditions can reduce or accentuate warming and affect the intensity, frequency and duration of extreme events. The magnitude and direction of these changes vary with location and season (*high confidence*). {Executive Summary Chapter 2, 2.3, 2.4, 2.5, 3.3}

A 5. Climate change creates additional stresses on land, exacerbating existing risks to livelihoods, biodiversity, human and ecosystem health, infrastructure, and food

¹ This assessment only includes CO₂, CH₄ and N₂O.

² Global food system in this report is defined as ‘all the elements (environment, people, inputs, processes, infrastructures, institutions, etc.) and activities that relate to the production, processing, distribution, preparation and consumption of food, and the output of these activities, including socioeconomic and environmental outcomes at the global level’. These emissions data are not directly comparable to the national inventories prepared according to the 2006 IPCC Guidelines for National Greenhouse Gas.

³ Land conditions encompass changes in land cover (e.g. deforestation, afforestation, urbanisation), in land use (e.g. irrigation), and in land state (e.g. degree of wetness, degree of greening, amount of snow, amount of permafrost)

systems (*high confidence*). Increasing impacts on land are projected under all future GHG emission scenarios (*high confidence*). Some regions will face higher risks, while some regions will face risks previously not anticipated (*high confidence*). Cascading risks with impacts on multiple systems and sectors also vary across regions (*high confidence*). {2.2, 3.5, 4.2, 4.4, 4.7, 5.1, 5.2, 5.8, 6.1, 7.2, 7.3, Cross-Chapter Box 9 in Chapter 6, Figure SPM.2}

A6. The level of risk posed by climate change depends both on the level of warming and on how population, consumption, production, technological development, and land management patterns evolve (*high confidence*). Pathways with higher demand for food, feed, and water, more resource-intensive consumption and production, and more limited technological improvements in agriculture yields result in higher risks from water scarcity in drylands, land degradation, and food insecurity (*high confidence*). {5.1.4, 5.2.3, 6.1.4, 7.2, Cross-Chapter Box 9 in Chapter 6, Figure SPM.2b}

B. Adaptation and mitigation response options

B 1. Many land-related responses that contribute to climate change adaptation and mitigation can also combat desertification and land degradation and enhance food security. The potential for land-related responses and the relative emphasis on adaptation and mitigation is context specific, including the adaptive capacities of communities and regions. While land-related response options can make important contributions to adaptation and mitigation, there are some barriers to adaptation and limits to their contribution to global mitigation. (*very high confidence*) {2.6, 4.8, 5.6, 6.1, 6.3, 6.4, Figure SPM.3}

B 2. Most of the response options assessed contribute positively to sustainable development and other societal goals (*high confidence*). Many response options can be applied without competing for land and have the potential to provide multiple co-benefits (*high confidence*). A further set of response options has the potential to reduce demand for land, thereby enhancing the potential for other response options to deliver across each of climate change adaptation and mitigation, combating desertification and land degradation, and enhancing food security (*high confidence*). {4.8, 6.2, 6.3.6, 6.4.3; Figure SPM.3}

B 3. Although most response options can be applied without competing for available land, some can increase demand for land conversion (*high confidence*). At the deployment scale of several GtCO₂yr⁻¹, this increased demand for land conversion could lead to adverse side effects for adaptation, desertification, land degradation and food security (*high confidence*). If applied on a limited share of total land and integrated into sustainably managed landscapes, there will be fewer adverse side-effects and some

positive co-benefits can be realised (*high confidence*). {4.5, 6.2, 6.4; Cross-Chapter Box 7 in Chapter 6; Figure SPM.3}

B 4. Many activities for combating desertification can contribute to climate change adaptation with mitigation co-benefits, as well as to halting biodiversity loss with sustainable development co-benefits to society (*high confidence*). Avoiding, reducing and reversing desertification would enhance soil fertility, increase carbon storage in soils and biomass, while benefitting agricultural productivity and food security (*high confidence*). Preventing desertification is preferable to attempting to restore degraded land due to the potential for residual risks and maladaptive outcomes (*high confidence*). {3.6.1, 3.6.2, 3.6.3, 3.6.4, 3.7.1, 3.7.2}

B 5. Sustainable land management⁴, including sustainable forest management⁵, can prevent and reduce land degradation, maintain land productivity, and sometimes reverse the adverse impacts of climate change on land degradation (*very high confidence*). It can also contribute to mitigation and adaptation (*high confidence*). Reducing and reversing land degradation, at scales from individual farms to entire watersheds, can provide cost effective, immediate, and long-term benefits to communities and support several Sustainable Development Goals (SDGs) with co-benefits for adaptation (*very high confidence*) and mitigation (*high confidence*). Even with implementation of sustainable land management, limits to adaptation can be exceeded in some situations (*medium confidence*). {1.3.2, 4.1.5, 4.8, Table 4.2}

B 6. Response options throughout the food system, from production to consumption, including food loss and waste, can be deployed and scaled up to advance adaptation and mitigation (*high confidence*). The total technical mitigation potential from crop and livestock activities, and agroforestry is estimated as 2.3-9.6 GtCO₂e.yr⁻¹ by 2050 (*medium confidence*). The total technical mitigation potential of dietary changes is estimated as 0.7-8 GtCO₂e.yr⁻¹ by 2050 (*medium confidence*). {5.3, 5.5, 5.6}

B 7. Future land use depends, in part, on the desired climate outcome and the portfolio of response options deployed (*high confidence*). All assessed modelled pathways that limit warming to 1.5°C or well below 2°C require land-based mitigation

⁴ Sustainable land management is defined in this report as the stewardship and use of land resources, including soils, water, animals and plants, to meet changing human needs, while simultaneously ensuring the long-term productive potential of these resources and the maintenance of their environmental functions. Examples of options include inter alia agroecology (including agroforestry), conservation agriculture and forestry practices, crop and forest species diversity, appropriate crop and forest rotations, organic farming, integrated pest management, the conservation of pollinators, rain water harvesting, range and pasture management, and precision agriculture systems.

⁵ Sustainable forest management is defined in this report as the stewardship and use of forests and forest lands in a way, and at a rate, that maintains their biodiversity, productivity, regeneration capacity, vitality, and their potential to fulfill now and in the future, relevant ecological, economic and social functions at local, national and global levels and that does not cause damage to other ecosystems.

and land-use change, with most including different combinations of reforestation, afforestation, reduced deforestation, and bioenergy (*high confidence*). A small number of modelled pathways achieve 1.5°C with reduced land conversion (*high confidence*) and, thus, reduced consequences for desertification, land degradation, and food security (*medium confidence*). {2.6, 6.4, 7.4, 7.6; Cross-Chapter Box 9 in Chapter 6; Figure SPM.4}

C. Enabling response options

C 1. Appropriate design of policies, institutions and governance systems at all scales can contribute to land-related adaptation and mitigation while facilitating the pursuit of climate-adaptive development pathways (*high confidence*). Mutually supportive climate and land policies have the potential to save resources, amplify social resilience, support ecological restoration, and foster engagement and collaboration between multiple stakeholders (*high confidence*). {Figure SPM.1, Figure SPM.2, Figure SPM.3; 3.6.2, 3.6.3, 4.8, 4.9.4, 5.7, 6.3, 6.4, 7.2.2, 7.3, 7.4, 7.4.7, 7.4.8, 7.5, 7.5.5, 7.5.6, 7.6.6; Cross-Chapter Box 10 in Chapter 7}

C2. Policies that operate across the food system, including those that reduce food loss and waste and influence dietary choices, enable more sustainable land-use management, enhanced food security and low emissions trajectories (*high confidence*). Such policies can contribute to climate change adaptation and mitigation, reduce land degradation, desertification and poverty as well as improve public health (*high confidence*). The adoption of sustainable land management and poverty eradication can be enabled by improving access to markets, securing land tenure, factoring environmental costs into food, making payments for ecosystem services, and enhancing local and community collective action (*high confidence*). {1.1.2, 1.2.1, 3.6.3, 4.7.1, 4.7.2, 4.8, 5.5, 6.4, 7.4.6, 7.6.5}

C 3. Acknowledging co-benefits and trade-offs when designing land and food policies can overcome barriers to implementation (*medium confidence*). Strengthened multilevel, hybrid and cross-sectoral governance, as well as policies developed and adopted in an iterative, coherent, adaptive and flexible manner can maximise co-benefits and minimise trade-offs, given that land management decisions are made from farm level to national scales, and both climate and land policies often range across multiple sectors, departments and agencies (*high confidence*). {Figure SPM.3; 4.8.5, 4.9, 5.6, 6.4, 7.3, 7.4.6, 7.4.8, 7.4.9, 7.5.6, 7.6.2}

C 4. The effectiveness of decision-making and governance is enhanced by the involvement of local stakeholders (particularly those most vulnerable to climate change including indigenous peoples and local communities, women, and the poor and marginalised) in the selection, evaluation, implementation and monitoring of policy instruments for land-based climate change adaptation and mitigation (*high confidence*). Integration across sectors and scales increases the chance of maximising co-benefits and minimising trade-offs (*medium confidence*). {1.4, 3.1, 3.6, 3.7, 4.8, 4.9, 5.1.3, Box 5.1, 7.4, 7.6}

D. Action in the near-term

D 1. Actions can be taken in the near-term, based on existing knowledge, to address desertification, land degradation and food security while supporting longer-term responses that enable adaptation and mitigation to climate change. These include actions to build individual and institutional capacity, accelerate knowledge transfer, enhance technology transfer and deployment, enable financial mechanisms, implement early warning systems, undertake risk management and address gaps in implementation and upscaling (*high confidence*). {3.6.1, 3.6.2, 3.7.2, 4.8, 5.3.3, 5.5, 5.6.4, 5.7, 6.2, 6.4, 7.3, 7.4.9, 7.6; Cross-Chapter Box 10 in Chapter 7}

D 2. Near-term action to address climate change adaptation and mitigation, desertification, land degradation and food security can bring social, ecological, economic and development co-benefits (*high confidence*). Co-benefits can contribute to poverty eradication and more resilient livelihoods for those who are vulnerable (*high confidence*). {3.4.2, 5.7, 7.5}

D 3. Rapid reductions in anthropogenic GHG emissions across all sectors following ambitious mitigation pathways reduce negative impacts of climate change on land ecosystems and food systems (*medium confidence*). Delaying climate mitigation and adaptation responses across sectors would lead to increasingly negative impacts on land and reduce the prospect of sustainable development (*medium confidence*). {Box SPM.1, Figure SPM.2, 2.5, 2.7, 5.2, 6.2, 6.4, 7.2, 7.3.1, 7.4.7, 7.4.8, 7.5.6; Cross-Chapter Box 9 in Chapter 6, Cross-Chapter Box 10 in Chapter 7}

Second Session of Working Group I, II and III

SRCCL Final Draft

| | SPM Page:Line | Chapter | Chapter Page:Line | Tricklebacks |
|----|--|---------|----------------------|--|
| 1 | 4:12 | 1 | 1:16 | simplify: "terrestrial" to "land" potential NPP; also need to adopt NPP definition in glossary |
| 2 | 4:27 | 1 | 2:28 | food waste and lost: need to check if we need to adjust confidence languag |
| 3 | A2.1 [8:7] | 1 | 1:25 | update land temperature increase and underlying time period, (abbreviation GMST) |
| 4 | A2.1 [8:7] | 1 | 8:25 | update land temperature increase and underlying time period, (abbreviation GMST) |
| 5 | A1.1 Footnote2 | 1 | 1:6 | drop range (75-85) to a single value of 75 trillion USD, revise sentence to reflect new SPM statement: "The world's terrestrial ecosystem services have been valued on an annual basis to be approximately equivalent to the total economic to the annual global Gross Domestic Product..." Adjust confidence statement. |
| 6 | A1.1 Footnote2 | 1 | 3:32 | drop range (75-85) to a single value of 75 trillion USD, revise statement to reflect new SPM statement: "The world's terrestrial ecosystem services have been valued on an annual basis to be approximately equivalent to the total economic to the annual global Gross Domestic Product..." |
| 7 | Fig SPM.1 | 1 | 5 | Replace with new SPM Figure, adjust Figure caption, and literature/data sources |
| 8 | | 1 | 10:8 | Replace (26.3-43-4 mio km2) by (32.0-42.5 mio km2) |
| 9 | Fig SPM.1 | 1 | 13 | Table 1:1 -- change area of Managed for timber... from 21 to 22% (was a rounding issue that doesn't add up). |
| 10 | Fig SPM.1 | 1 | 13:17 | In 1.2.2.3 add fibre and timber numbers (if still in Figure) |
| 11 | Fig SPM.1 | 1 | 12:16 | change to inland wetland area |
| 12 | Fig SPM.1 | 1 | panel B | change emissions to 23% (from 22%) in response to the changed emission estimates in Table SPM.1 |
| 13 | A1.5 | 1 | 15:16 | Change to word similar to the revised SPM: In 2015, about 500 (360-620) million people lived within areas which experienced desertification between 1980s and 2000s |
| 14 | SPM.4 | 1 | 21:42 | Change to ...not easily commensurate... |
| 15 | B5 HeadlineS. Footnote 33 | 1 | 31:4 | cross-check definition of "sustainable land management" with glossary, (following changes in B5 footnote 33) |
| 16 | table SPM1 and text SPM A3 headline] | 2 | 4:16 | change 22 % to 23% |

Tricklebacks

| | SPM Page:Line | Chapter | Chapter Page:Line | |
|----|------------------------------|---------|-------------------|---|
| 17 | table SPM1 | 2 | 4: 23-24 | Change "provided a global net removal from 2008 to 2017 according to models, (-6.2 ± 3.7 GtCO ₂ yr ⁻¹ , medium 23 confidence)." change date range and number to "provided a global net removal from 2007 to 2016 according to models, (-6.0 ± 3.7 GtCO ₂ yr ⁻¹ , likely range)". |
| 18 | table SPM1 and text SPM A3.1 | 2 | 4:25 | change number from 5.5 to 5.2 ± 2.6 GtCO ₂ yr ⁻¹ |
| 19 | table SPM1 and text SPM A3.1 | 2 | 4:27 | delele "medium confidence"and change "likely" to " likely range " |
| 20 | table SPM1 and text SPM A3.2 | 2 | 4:28 | replace -11.7 ± 2.6. with. -11.2 ± 2.6 and change " <i>likely</i> " to " likely range " |
| 21 | SPM A3,3 | 2 | 4:32 to 41 | Replace first sentence of headline with SPM text : "Global models and national GHG inventories use different methods to estimate anthropogenic CO ₂ emissions and removals for the land sector." Replace forst part of second headline sentence with bold text from SPM. : "Consideration of differences in methods can enhance understanding of land sector net emission such as under the Paris Agreement's global stocktake (medium confidence)." these two new sentences reflecting the SPM should be the new bold headlien statement. Then replace the rest of this ES bullet entirely with SPM text "Both models and inventories produce estimates that are in close agreement for land-use change involving forest (e.g., deforestation, afforestation), and differ for managed forest. Global models consider as managed forest those lands that were subject to harvest whereas, consistent with IPCC guidelines, national GHG inventories define managed forest more broadly. On this larger area, inventories can also consider the natural response of land to human-induced environmental changes as anthropogenic, while the global model approach {Table SPM.1} treats this response as part of the non-anthropogenic sink. For illustration, from 2005 to 2014, the sum of the national GHG inventories net emission estimates is 0.1±1.0 GtCO ₂ yr ⁻¹ , while the mean of two global bookkeeping models is 5.1±2.6 GtCO ₂ yr ⁻¹ (likely range). |
| 22 | table SPM1 | 2 | 4:51 | replace % and year range with 44% and 2006-17 |
| 23 | table SPM1 | 2 | 28:36 | replace number with 23% |
| 24 | table SPM1 | 2 | 28:37 | replace time period with 2007-2016 |
| 25 | table SPM1 | 2 | 28:43 | this entire table will be prepalced with new SPM table 1, panel A and associated footnotes |
| 26 | table SPM1 | 2 | 29:27 to 28 | change number and add confidence 6.0 ± 2.0 GtCO ₂ yr ⁻¹ , (likely range),. change time period to 2007-2016 |

Tricklebacks

| | SPM Page:Line | Chapter | Chapter Page:Line | |
|----|-----------------------------------|---------|-------------------|--|
| 27 | table SPM1, bullets A3.1 and A3.2 | 2 | 30:22 to 31:1 | table 2.3 needs a new column added with the time period 2007-2016 to be consistent with table SPM1m and bullets, and as an updated version to the GCP database was used to calculate table SPM1, the consistent version of the database will need to be used to recalculate numbers in table 2.3, although most will probably stay the same, a table legend will need to include information that the updated database is used rather than the original numbers in the paper Le Quere et al 2018 |
| 28 | table SPM1 and text SPM A3.1 | 2 | 31:3 | update number and time period and add confidence statement: 5.2 ± 2.6 GtCO ₂ yr ⁻¹ (likely range) 2007-2016 |
| 29 | table SPM1 footnote 2 | 2 | 31:6 to 7 | Delete "including agriculture, grasslands and shrub" and replace with "as well as peatland drainage and burning" |
| 30 | A3.3 | 2 | 32:24 | Delete low confidence and use (<i>likely</i> range), as consistent with bullet |
| 31 | A3.3 | 2 | 32:22 | add confidence statement after number (<i>likely</i> range) |
| 32 | table SPM1, bullet A3.2 | 2 | 36:5 | change number, time period and uncertainty: 11.2 ± 2.6 Gt CO ₂ yr ⁻¹ (likely range) during 2007-2016 |
| 33 | table SPM1, bullet A3.1 | 2 | 36:41 | change number, time period and uncertainty 5.2 ± 2.6 GtCO ₂ yr ⁻¹ (likely range) over the period 2007-2016 |
| 34 | table SPM1, bullet A3.2 | 2 | 37:2 | change number, time period and uncertainty: 11.2 ± 2.6 Gt CO ₂ yr ⁻¹ (likely range) during 2007-2016 |
| 35 | | 2 | TS.2 | Replace second sentence in the technical summary TS.2 with the values for 2006-2015 period for land and GMST temperature increase: "According to the single longest and most extensive dataset, from 1850-1900 to 2006-2015 mean land surface air temperature has increased by 1.53°C (very likely range from 1.38°C to 1.68°C) while GMST increased by 0.87°C (likely range from 0.75°C to 0.99°C). For the 1881–2018 period..." |
| 36 | SPM A2.1 | 2 | p.16,l.1 | replace "the 1999–2018 period was 1.52°C, 1 (1.39°C–1.66°C; 95% confidence)." with "the 2006-2015 period was 1.53°C, (1.38°C–1.68°C; 95% confidence), while GMST increase for the same period was by 0.87°C (0.75°C -0.99°C; 90% confidence) (IPCC, 2018: Summary for Policymakers). |
| 37 | | 2 | | add reference for the SR1.5 SPM in chapter 2: IPCC, 2018: Summary for Policymakers. In: Global Warming of 1.5°C. An IPCC Special Report on the impacts of global warming of 1.5°C above pre-industrial levels and related global greenhouse gas emission pathways, in the context of strengthening the global response to the threat of climate change, sustainable development, and efforts to eradicate poverty [Masson-Delmotte, V., P. Zhai, H.-O. Pörtner, D. Roberts, J. Skea, P.R. Shukla, A. Pirani, W. Moufouma-Okia, C. Péan, R. Pidcock, S. Connors, J.B.R. Matthews, Y. Chen, X. Zhou, M.I. Gomis, E. Lonnoy, T. Maycock, M. Tignor, and T. Waterfield (eds.)] |
| 38 | | 2 | p16. l.6-7 | Change table 2.1 name to "Table 2.1 Increases in land surface air temperature (LSAT) from preindustrial period and the late 19th century to present " |
| 39 | | 2 | p. 16 | Change table 2.1, including the labels on the first column and adding a new row for temp increases from preindustrial to 2006-2015 |

Tricklebacks

| | SPM Page:Line | Chapter | Chapter Page:Line | |
|----|---------------|---------|--|--|
| 40 | A2.2 | 3 | 3:11; 11:4 | "Middle East" to be replaced by " West Asia " for drought [take note as it is somehow contradict with A1.5 approved bullet] |
| 41 | B4.1 | 3 | 5:16; 50:17 | combating desertification are " site and regionally specific " |
| 42 | B4.1 | 3 | 4:38-39; 9:17; 61:29 | drought-resilient native plants (drought- tolerant crops - initial version in the underlying report) ----> drought-resilient ecological appropriate plants |
| 43 | B4.1 | 3 | 32:24 | drought-resilient native plants (drought- adapted crops - initial version in the underlying report) ----> drought-resilient ecological appropriate plants |
| 44 | B4.2 | 3 | 69:3; 71:33; 71:43; 77:6-7 | "native tree species" ----> "native and other climate resilient tree species" |
| 45 | B4.4 | 3 | 5:23-24; 12:39; 33:25; 44:1; 58:3 | poverty reduction ----> poverty eradication (or eradicating poverty) |
| 46 | B4.4 | 3 | 5:25; 53:15 | LDN ----> LDN (including avoiding, reducing and reversing land degradation) |
| 47 | A1.5 | 3 | 3:10-11 | The highest numbers of people affected are in South and East Asia, North Africa and Middle East (low confidence) ----> The highest numbers of people affected are in South and East Asia, the circum Sahara region including North Africa, and the Middle East including the Arabian peninsula (low confidence). Other dryland regions have also experienced desertification. [take note as it is somehow contradict with A2.2 approved bullet] |
| 48 | B4.5 | 3 | 3.1; 3.7.4 & 3.9 (sections) | To add confidence level and missing words: "Even when solutions are available, social, economic and institutional constraints could pose barriers to their implementation (medium confidence). |
| 49 | B4.6 | 3 | X | Developing, enabling and promoting access to cleaner energy sources and technologies can contribute to adaptation and mitigating climate change and combating desertification and forest degradation through decreasing the use of traditional biomass for energy while increasing the diversity of energy supply (medium confidence). [Throughout the entire Ch3 report, we use renewable energy] |
| 50 | B4.6 | 3 | X | This can have socioeconomic and health benefits, especially for women and children . (<i>high confidence</i>). |
| 51 | B4 | 3 | 6:6; 6:8; 10:13; 12:39 |climate change adaptation and mitigation... ----> ...climate change adaptation and with mitigation co-benefits |
| 52 | 8:29-31 | 4 | 4:25 | Add the following sentence between "confidence)" and "Global" In some areas sea level rise has exacerbated coastal erosion (medium confidence). |

Tricklebacks

| | SPM Page:Line | Chapter | Chapter Page:Line | |
|----|----------------------------|---------|----------------------|---|
| 53 | 4:28 | 4 | 5:27 | change 1-6 Mkm2 to 10-60 MKm2 We spotted a decimal error. It is correct in the chapter text, but was incorrectly inserted into the ES. |
| 54 | 8:34 (A2.7) | 5 | 5:16 | Add "over recent decades" after "regions". |
| 55 | 8:34 (A2.7) | 5 | 5:33 | Add "in Africa" after "pastoral systems" |
| 56 | 4:23 (new A1.4) | 5 | 7:9 | Change "amount to a third of global food production" to "amount to 25-30% of total food produced". Change to "medium confidence" |
| 57 | 4:23 (new A1.4) | 5 | 80:21 | Change "amount to a third of global food production" to "amount to 25-30% of total food produced". Change to "medium confidence" |
| 58 | TableSPM.1, Footnote 10 | 5 | 7:11 | Update "total GHG emissions from food systems" to "total anthropogenic GHG emissions" |
| 59 | TableSPM.1, Footnote 10 | 5 | 61:23 | Update "total GHG emissions from agriculture and land use" to "total anthropogenic GHG emissions" |
| 60 | TableSPM.1, Footnote 10 | 5 | 80:34 | Update "emissions of the entire food system" to "emissions" |
| 61 | TableSPM.1 | 5 | 60:29 | Update "6.2 ± 1.9" to "6.2 ± 1.4" |
| 62 | TableSPM.1 | 5 | 64:37 | Update "15%" to "10%" |
| 63 | TableSPM.1 | 5 | 60:30 | Update "4.1 ± 1.2" to "4.0 ± 1.2" |
| 64 | TableSPM.1 | 5 | 60:31 | Update "8.3 ± 2.3" to "8.0 ± 2.0" |
| 65 | TableSPM.1 | 5 | 60:31 | Update "2.1 ± 0.6" to "2.2 ± 0.7" |
| 66 | TableSPM.1 | 5 | 60:34 | Update "4.8 ± 2.4" to "4.9 ± 2.5" |
| 67 | TableSPM.1 | 5 | 60:38 | Update "11.0 ± 3.1" to "11.1 ± 2.9" |
| 68 | TableSPM.1 | 5 | 61:25 (Table 5.4) | Update first row "6.2 ± 1.9" to "6.2 ± 1.4" |
| 69 | TableSPM.1 | 5 | 61:25 (Table 5.4) | Update second row "4.8 ± 2.4" to "4.9 ± 2.5" |
| 70 | TableSPM.1 | 5 | 61:25 (Table 5.4) | Update third row "3.8 ± 1.3" to "2.4 - 4.8" |
| 71 | TableSPM.1 | 5 | 61:25 (Table 5.4) | Update fourth row "14.8 ± 3.4" to "10.7 - 19.1" |
| 72 | TableSPM.1 | 5 | 61:20 | Update "2.5" to "2.4" |
| 73 | TableSPM.1 | 5 | 61:25 (Table 5.4) | Update first row "10-12%" to "10-14%" |

Tricklebacks

| | SPM Page:Line | Chapter | Chapter Page:Line | |
|----|---------------|---------|-------------------|--|
| 74 | TableSPM.1 | 5 | 61:25 (Table 5.4) | Update second row "8-10%" to "5-14%" |
| 75 | TableSPM.1 | 5 | 61:25 (Table 5.4) | Update fourth row "25-30%" to "21-37%" |
| 76 | TableSPM.1 | 5 | 61:29 | Update "51 Gt CO ₂ -eq yr ⁻¹ " to "52 Gt CO ₂ -eq yr ⁻¹ " |
| 77 | TableSPM.1 | 5 | 61:30 | Delete "GWP values used are those , and by". Sentence would read: "...(See Chapter 2), using GWP..." |
| 78 | TableSPM.1 | 5 | 61:28 | Update "a) FAOSTAT (2018) and US EPA (See also Chapter 2) and" to "a) FAOSTAT (2018) and US EPA (2012) (using square root of sum of squares of standard deviations for uncertainty ranges; see also Chapter 2) and" |
| 79 | TableSPM.1 | 5 | 61:28 | Update "b) Garnett (2011) and Poore and Nemecek (2018)" to "b) estimates from Fishedick et al. (2014) and Poore and Nemecek (2018)" |
| 80 | TableSPM.1 | 5 | 6:16 | Update "25-30%" to "21-37%" |
| 81 | TableSPM.1 | 5 | 6:18 | Update "10-12%" to "10-14%" |
| 82 | TableSPM.1 | 5 | 6:19 | Update "8-10%" to "5-14%" |
| 83 | TableSPM.1 | 5 | 6:24 | Update "8.3 ± 2.3" to "8.0 ± 2.0" |
| 84 | TableSPM.1 | 5 | 6:25 | Update "4.8 ± 2.4" to "4.9 ± 2.5" |
| 85 | TableSPM.1 | 5 | 6:27 | Update "11.0 ± 3.1" to "11.1 ± 2.9" |
| 86 | TableSPM.1 | 5 | 6:27 | Update "6.2 ± 1.9" to "6.2 ± 1.4" |
| 87 | 21:42 (B6) | 5 | 6:34 | Update "Total mitigation potential of crop and livestock activities is estimated as 1.5–4.0 GtCO ₂ -eq yr ⁻¹ by 2030 at prices ranging from 20-100 USD/tCO ₂ eq (high confidence)" to "Total technical mitigation potential from crop and livestock activities and agroforestry is estimated as 2.3 - 9.6 GtCO ₂ -eq yr ⁻¹ by 2050 (medium confidence)" |
| 88 | 21:44 (B6) | 5 | 6:47 | Update "Total mitigation potential of dietary changes is estimated as 1.8-3.4 GtCO ₂ eq yr ⁻¹ by 2050 at prices ranging from 20-100 USD/tCO ₂ (medium confidence)" to "Total technical mitigation potential of dietary changes is estimated as 0.7 - 8.0 GtCO ₂ eq yr ⁻¹ by 2050 (medium confidence)" |
| 89 | B1 | 6 | All | "Poverty reduction" should be changed to "poverty eradication" throughout the chapter |
| 90 | B2 | 6 | All | "Improved forest management" should be changed to "forest management" throughout the chapter |
| 91 | Figure SPM.3 | 6 | 97:4 | The red cells in the Desertification and Land degradation column of the bioenergy and BECCS row should be removed. |

Tricklebacks

| | SPM Page:Line | Chapter | Chapter Page:Line | |
|-----|---------------|---------|-----------------------------------|---|
| 92 | Figure SPM.3 | 6 | 97:4 | The text in the last column of the bioenergy and BECCS row should be changed from "Desertification: Moderate adverse impacts of large-scale bioenergy and BECCS through increased pressure on land (Table 6.33). Land degradation: Large adverse impacts of large-scale bioenergy and BECCS through increased pressure on land (Table 6.41)." to "Desertification: Up to 15 million km ² of additional land is required in 2100 in 2°C scenarios, which will increase pressure for desertification and land degradation {6.4.3.1.5; 6.4.4.1.5}. This defines the maximum area potentially impacted, though the actual area affected by this additional pressure is not easily quantified. Land degradation: Up to 15 million km ² of additional land is required in 2100 in 2°C scenarios, which will increase pressure for desertification and land degradation {6.4.3.1.5; 6.4.4.1.5}. This defines the maximum area potentially impacted, though the actual area affected by this additional pressure is not easily quantified." |
| 93 | | 6 | 5:25, 10:25, 115:4, 140:9, 141:13 | "Indigenous peoples and local communities' was phrasing adopted in SPM |
| 94 | | 6 | 5:25 | add producers before consumers |
| 95 | Box SPM1 | 6 | 142:34 | "Illustrative Futures" should say "Shared Socioeconomic Pathways" |
| 96 | Box SPM1 | 6 | 142:34 | First sentence of section "Illustrative Futures" should say "The five SSPs" and not "The three illustrative futures" |
| 97 | Box SPM1 | 6 | 142:34 | Text in section "Illustrative Futures", starting with "SSP1" should be replaced with final SSP description bullets from Box SPM.1 (but retaining the references in the current paragraph) |
| 98 | Figure SPM.4 | 6 | 142:34 | Figure should be replaced with the figures in panel A of the final Figure SPM.4 |
| 99 | Figure SPM.4 | 6 | 142:34 | "Policies in Illustrative Futures" should say "Policies in these Pathways" |
| 100 | Figure SPM.4 | 6 | 142:34 | The section on "Land use and land cover change" should use the titles and text from the panel labels in the final Figure SPM.4a |
| 101 | Figure SPM.4 | 6 | 142:34 | Table should be replaced with the table in panel B of the final Figure SPM.4 |
| 102 | Figure SPM.4 | 6 | 142:34 | Caption for figure should be edited to match the final caption of panel A in Figure SPM.4a (but last sentence with the citation should be retained) |
| 103 | B1.1 | 7 | 100 (Table) | Policy/Programmes/Instriestponse options are in Table 7.5. Please add 7.6.6. in the list of reference. |
| 104 | B1.2 | 7 | 100 (Table) | Policy/Programmes/Instriestponse options are in Table 7.5. Please add 7.6.6. in the list of reference. |
| 105 | BOX A6 | 7 | 101:1 (Box) | Cross-Chapter Box 9 exists in both Chpater 6 and 7. Suggest to specify the location of Cross-Chapter Box 9 at the end of Box A6. Namely, Cross-Chapter Box 9 in Chapter 6 and 7. |
| 106 | B1.3 | 7 | 30 (Box) | Cross-Chapter Box 10 in Chapter 7 and Table 7.5 in Chapter 7 supporting policies and economic dimensions. So, please add Cross-Box 10 and 7.6.6. in the reference list. |

Tricklebacks

| | SPM Page:Line | Chapter | Chapter Page:Line | |
|-----|----------------|----------|-------------------|---|
| 107 | | 7 | 100 (Table) | Add sentence regarding other latitudes as in agreed SPM precipitation bullet and reference to new TS precipitation figure, which consists of a six panel figure (3 datasets by 2 periods) |
| 108 | B1.1. | 7 | 73(Box) | Forest conservation instruments are addressed in 7.5.6.4 and Table 7.5 in 7.6.6. |
| 109 | B5 | 7 | 30 (Box) | Policy/Programmes/Instruments for Sustainable land management are in Table 7.5. So, please add 7.6.6. in the list of reference. |
| 110 | C1.1. | 7 | | Land-use zoning, spatial planning, standards, payment for ecosystem services etc are covered slightly over the multiple subsection of the chapter. Please replace it 7.4, 7.5.6, 7.6. |
| 111 | C1.3. | 7 | 69:37 | 7.5.5. covers LDN mostly, please replace 7.4.5 with 7.5.5 |
| 112 | C2.2 | 7 | | 7.5.4, Cross-Chapter Box 10 in Chapter 7 |
| 113 | C3.4. | 7 | 101 | Maladaptation is address in Cross-Chapter Box 9 in 7.6.6. |
| 114 | C3.4. | 7 | 111 | Maladaptation is address in 7.6.6.5. |
| 115 | C3.4. | 7 | 141(box) | Frequently ased questions box addressed maladaptation breifly. |
| 116 | D1 | 7 | | 7.4, 11 7.5.9, 7.7; Cross-Chapter Box 10 in Chapter 7 |
| 117 | D2.1. | 7 | | 7.3, 7.4, 7.5, 7.6; Cross-Chapter Box 12 in Chapter 7 |
| 118 | 4:Footnote (1) | Glossary | 44 | Update definitions of net primary production (NPP) in the glossary, as agreed in SPM, to: 'the amount of carbon accumulated through photosynthesis minus the amount lost by plant respiration over a specied time period that would prevail in the absence of land use. |
| 119 | Footnote | Glossary | 55 | Define vegetation greening: An increase in photosynthetically active plant biomass which is inferred from satellite observations. |
| 120 | Footnote | Glossary | 14 | Define CO2 fertilisation: The enhancement of plant growth as a result of increased atmospheric carbon dioxide (CO2) concentration. The magnitude of CO2 fertilisation depends on nutrients and water availability. |
| 121 | Footnote | Glossary | 55 | Define vegetation browning: A decrease in photosynthetically active plant biomass which is inferred from satellite observations. |
| 122 | | Glossary | 4 | Define FOLU as a sub-term under AFOLU: FOLU (Forestry and Other Land Use)—also referred to as LULUCF (Land Use, Land-Use Change, and Forestry)—is the subset of AFOLU emissions and removals of greenhouse gases (GHGs) resulting from direct human-induced land use, land-use change, and forestry activities excluding agricultural emissions. {WGIII} |
| 123 | SPM A2.1 | TS | p. 10, l. 9:11 | Replace second sentence and the begining of the next sentence in the technical summary TS.2 with the values for 2006-2015 period for land and GMST temperature increase: "According to the single longest and most extensive dataset, from 1850-1900 to 2006-2015 mean land surface air temperature has increased by 1.53°C (very likely range from 1.38°C to 1.68°C) while GMST increased by 0.87°C (likely range from 0.75°C to 0.99°C). For the 1881–2018 period..." |
| 124 | SPM A2.1 | TS | p.10 | replace figure TS.4 with figure 2.2 from chapter 2, page 2-14. The paragraph refers to temperature changes but the figure shows schematic diagram. |
| 125 | | TS | p.13 | replace figure TS.5 with figure from chapter 2.4, page-2-30. The TS uses an earlier version of the figure with incorrect labels and outdated graphics |

Potential global contribution of response options to mitigation, adaptation, combating desertification and land degradation, and enhancing food security

Panel A shows response options that can be implemented without or with limited competition for land, including some that have the potential to reduce the demand for land. Co-benefits and adverse side effects are shown quantitatively based on the high end of the range of potentials assessed. Magnitudes of contributions are categorised using thresholds for positive or negative impacts. Letters within the cells indicate confidence in the magnitude of the impact relative to the thresholds used (see legend). Confidence in the direction of change is generally higher.

| Response options based on land management | | Mitigation | Adaptation | Desertification | Land Degradation | Food Security | Cost |
|--|--|------------|------------|-----------------|------------------|---------------|------|
| Agriculture | Increased food productivity | L | M | L | M | H | --- |
| | Agro-forestry | M | M | M | M | L | ● |
| | Improved cropland management | M | L | L | L | L | ●● |
| | Improved livestock management | M | L | L | L | L | ●●● |
| | Agricultural diversification | L | L | L | M | L | ● |
| | Improved grazing land management | M | L | L | L | L | --- |
| | Integrated water management | L | L | L | L | L | ●● |
| | Reduced grassland conversion to cropland | L | --- | L | L | -L | ● |
| Forests | Forest management | M | L | L | L | L | ●● |
| | Reduced deforestation and forest degradation | H | L | L | L | L | ●● |
| Soils | Increased soil organic carbon content | H | L | M | M | L | ●● |
| | Reduced soil erosion | ↔ L | L | M | M | L | ●● |
| | Reduced soil salinization | --- | L | L | L | L | ●● |
| | Reduced soil compaction | --- | L | --- | L | L | ● |
| Other ecosystems | Fire management | M | M | M | M | L | ● |
| | Reduced landslides and natural hazards | L | L | L | L | L | --- |
| | Reduced pollution including acidification | ↔ M | M | L | L | L | --- |
| | Restoration & reduced conversion of coastal wetlands | M | L | M | M | ↔ L | --- |
| | Restoration & reduced conversion of peatlands | M | --- | na | M | -L | ● |
| Response options based on value chain management | | | | | | | |
| Demand | Reduced post-harvest losses | H | M | L | L | H | --- |
| | Dietary change | H | --- | L | H | H | --- |
| | Reduced food waste (consumer or retailer) | H | --- | L | M | M | --- |
| Supply | Sustainable sourcing | --- | L | --- | L | L | --- |
| | Improved food processing and retailing | L | L | --- | --- | L | --- |
| | Improved energy use in food systems | L | L | --- | --- | L | --- |
| Response options based on risk management | | | | | | | |
| Risk | Livelihood diversification | --- | L | --- | L | L | --- |
| | Management of urban sprawl | --- | L | L | M | L | --- |
| | Risk sharing instruments | ↔ L | L | --- | ↔ L | L | ●● |

Options shown are those for which data are available to assess global potential for three or more land challenges. The magnitudes are assessed independently for each option and are not additive.

Key for criteria used to define magnitude of impact of each integrated response option

| | Mitigation Gt CO ₂ -eq yr ⁻¹ | Adaptation Million people | Desertification Million km ² | Land Degradation Million km ² | Food Security Million people |
|-----------------|---|------------------------------|--|---|---------------------------------|
| Positive | | | | | |
| Large | More than 3 | Positive for more than 25 | Positive for more than 3 | Positive for more than 3 | Positive for more than 100 |
| Moderate | 0.3 to 3 | 1 to 25 | 0.5 to 3 | 0.5 to 3 | 1 to 100 |
| Small | Less than 0.3 | Less than 1 | Less than 0.5 | Less than 0.5 | Less than 1 |
| Negligible | No effect | No effect | No effect | No effect | No effect |
| Negative | | | | | |
| Small | Less than -0.3 | Less than 1 | Less than 0.5 | Less than 0.5 | Less than 1 |
| Moderate | -0.3 to -3 | 1 to 25 | 0.5 to 3 | 0.5 to 3 | 1 to 100 |
| Large | More than -3 | Negative for more than 25 | Negative for more than 3 | Negative for more than 3 | Negative for more than 100 |

↔ Variable: Can be positive or negative --- no data na not applicable

Confidence level

Indicates confidence in the estimate of magnitude category.

H High confidence
M Medium confidence
L Low confidence

Cost range

See technical caption for cost ranges in US\$ tCO₂e⁻¹ or US\$ ha⁻¹.

●●● High cost
●● Medium cost
● Low cost
--- no data

Potential global contribution of response options to mitigation, adaptation, combating desertification and land degradation, and enhancing food security

Panel B shows response options that rely on additional land-use change and could have implications across three or more land challenges under different implementation contexts. For each option, the first row (high level implementation) shows a quantitative assessment (as in Panel A) of implications for global implementation at scales delivering CO₂ removals of more than 3 GtCO₂ yr⁻¹ using the magnitude thresholds shown in Panel A. The red hatched cells indicate an increasing pressure but unquantified impact. For each option, the second row (best practice implementation) shows qualitative estimates of impact if implemented using best practices in appropriately managed landscape systems that allow for efficient and sustainable resource use and supported by appropriate governance mechanisms. In these qualitative assessments, green indicates a positive impact, grey indicates a neutral interaction.

Bioenergy and BECCS



High level: Impacts on adaptation, desertification, land degradation and food security are maximum potential impacts, assuming carbon dioxide removal by BECCS at a scale of 11.3 GtCO₂ yr⁻¹ in 2050, and noting that bioenergy without CCS can also achieve emissions reductions of up to several GtCO₂ yr⁻¹ when it is a low carbon energy source {2.7.1.5; 6.4.1.1.5}. Studies linking bioenergy to food security estimate an increase in the population at risk of hunger to up to 150 million people at this level of implementation {6.4.5.1.5}. The red hatched cells for desertification and land degradation indicate that while up to 15 million km² of additional land is required in 2100 in 2°C scenarios which will increase pressure for desertification and land degradation, the actual area affected by this additional pressure is not easily quantified {6.4.3.1.5; 6.4.4.1.5}.



Best practice: The sign and magnitude of the effects of bioenergy and BECCS depends on the scale of deployment, the type of bioenergy feedstock, which other response options are included, and where bioenergy is grown (including prior land use and indirect land use change emissions). For example, limiting bioenergy production to marginal lands or abandoned cropland would have negligible effects on biodiversity, food security, and potentially co-benefits for land degradation; however, the benefits for mitigation could also be smaller. {Table 6.58}

Reforestation and forest restoration



High level: Impacts on adaptation, desertification, land degradation and food security are maximum potential impacts assuming implementation of reforestation and forest restoration (partly overlapping with afforestation) at a scale of 10.1 GtCO₂ yr⁻¹ removal {6.4.1.1.2}. Large-scale afforestation could cause increases in food prices of 80% by 2050, and more general mitigation measures in the AFOLU sector can translate into a rise in undernourishment of 80–300 million people; the impact of reforestation is lower {6.4.5.1.2}.



Best practice: There are co-benefits of reforestation and forest restoration in previously forested areas, assuming small scale deployment using native species and involving local stakeholders to provide a safety net for food security. Examples of sustainable implementation include, but are not limited to, reducing illegal logging and halting illegal forest loss in protected areas, reforesting and restoring forests in degraded and desertified lands {Box6.1C; Table 6.6}.

Afforestation



High level: Impacts on adaptation, desertification, land degradation and food security are maximum potential impacts assuming implementation of afforestation (partly overlapping with reforestation and forest restoration) at a scale of 8.9 GtCO₂ yr⁻¹ removal {6.4.1.1.2}. Large-scale afforestation could cause increases in food prices of 80% by 2050, and more general mitigation measures in the AFOLU sector can translate into a rise in undernourishment of 80–300 million people {6.4.5.1.2}.



Best practice: Afforestation is used to prevent desertification and to tackle land degradation. Forested land also offers benefits in terms of food supply, especially when forest is established on degraded land, mangroves, and other land that cannot be used for agriculture. For example, food from forests represents a safety-net during times of food and income insecurity {6.4.5.1.2}.

Biochar addition to soil



High level: Impacts on adaptation, desertification, land degradation and food security are maximum potential impacts assuming implementation of afforestation at a scale of 6.6 GtCO₂ yr⁻¹ removal {6.4.1.1.3}. Dedicated energy crops required for feedstock production could occupy 0.4–2.6 Mkm² of land, equivalent to around 20% of the global cropland area, which could potentially have a large effect on food security for up to 100 million people {6.4.5.1.3}.



Best practice: When applied to land, biochar could provide moderate benefits for food security by improving yields by 25% in the tropics, but with more limited impacts in temperate regions, or through improved water holding capacity and nutrient use efficiency. Abandoned cropland could be used to supply biomass for biochar, thus avoiding competition with food production; 5-9 Mkm² of land is estimated to be available for biomass production without compromising food security and biodiversity, considering marginal and degraded land and land released by pasture intensification {6.4.5.1.3}.

A. Pathways linking socioeconomic development, mitigation responses and land

Socioeconomic development and land management influence the evolution of the land system including the relative amount of land allocated to **CROPLAND**, **PASTURE**, **BIOENERGY CROPLAND**, **FOREST**, and **NATURAL LAND**. The lines show the median across Integrated Assessment Models (IAMs) for three alternative shared socioeconomic pathways (**SSP1**, **SSP2** and **SSP5** at **RCP1.9**); shaded areas show the range across models. Note that pathways illustrate the effects of climate change mitigation but not those of climate change impacts or adaptation.

A. Sustainability-focused (SSP1)

Sustainability in land management, agricultural intensification, production and consumption patterns result in reduced need for agricultural land, despite increases in per capita food consumption. This land can instead be used for reforestation, afforestation, and bioenergy.

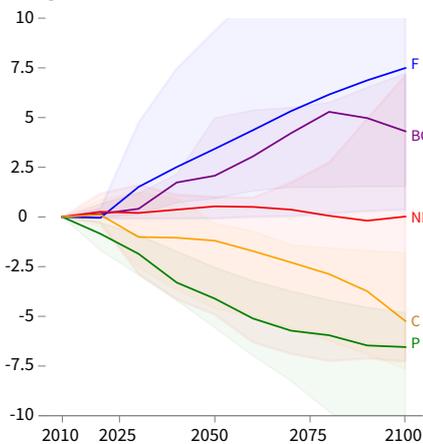
B. Middle of the road (SSP2)

Societal as well as technological development follows historical patterns. Increased demand for land mitigation options such as bioenergy, reduced deforestation or afforestation decreases availability of agricultural land for food, feed and fibre.

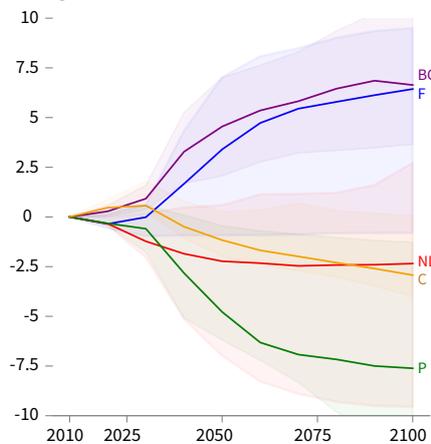
C. Resource intensive (SSP5)

Resource-intensive production and consumption patterns, results in high baseline emissions. Mitigation focuses on technological solutions including substantial bioenergy and BECCS. Intensification and competing land uses contribute to declines in agricultural land.

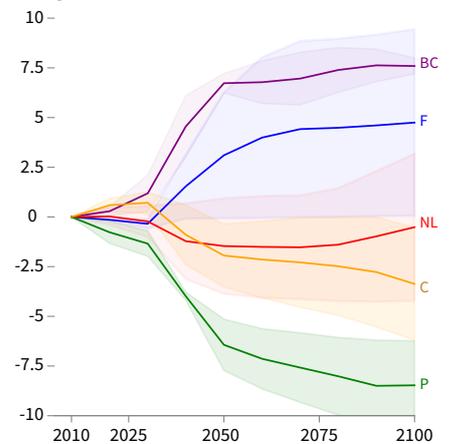
SSP1 Sustainability-focused
Change in Land from 2010 (Mkm²)



SSP2 Middle of the road
Change in Land from 2010 (Mkm²)



SSP5 Resource intensive
Change in Land from 2010 (Mkm²)



■ CROPLAND ■ PASTURE ■ BIOENERGY CROPLAND ■ FOREST ■ NATURAL LAND

B. Land use and land cover change in the SSPs

| Quantitative indicators for the SSPs | Count of models included* | Change in Natural Land from 2010 Mkm ² | Change in Bioenergy Cropland from 2010 Mkm ² | Change in Cropland from 2010 Mkm ² | Change in Forest from 2010 Mkm ² | Change in Pasture from 2010 Mkm ² | |
|--------------------------------------|---------------------------|---|---|---|---|--|--------------------|
| SSP1 | RCP1.9 in 2050 | 5/5 | 0.5 (-4.9, 1) | 2.1 (0.9, 5) | -1.2 (-4.6, -0.3) | 3.4 (-0.1, 9.4) | -4.1 (-5.6, -2.5) |
| | ↳ 2100 | | 0 (-7.3, 7.1) | 4.3 (1.5, 7.2) | -5.2 (-7.6, -1.8) | 7.5 (0.4, 15.8) | -6.5 (-12.2, -4.8) |
| | RCP2.6 in 2050 | 5/5 | -0.9 (-2.2, 1.5) | 1.3 (0.4, 1.9) | -1 (-4.7, 1) | 2.6 (-0.1, 8.4) | -3 (-4, -2.4) |
| | ↳ 2100 | | 0.2 (-3.5, 1.1) | 5.1 (1.6, 6.3) | -3.2 (-7.7, -1.8) | 6.6 (-0.1, 10.5) | -5.5 (-9.9, -4.2) |
| | RCP4.5 in 2050 | 5/5 | 0.5 (-1, 1.7) | 0.8 (0.5, 1.3) | 0.1 (-3.2, 1.5) | 0.6 (-0.7, 4.2) | -2.4 (-3.3, -0.9) |
| | ↳ 2100 | | 1.8 (-1.7, 6) | 1.9 (1.4, 3.7) | -2.3 (-6.4, -1.6) | 3.9 (0.2, 8.8) | -4.6 (-7.3, -2.7) |
| SSP2 | Baseline in 2050 | 5/5 | 0.3 (-1.1, 1.8) | 0.5 (0.2, 1.4) | 0.2 (-1.6, 1.9) | -0.1 (-0.8, 1.1) | -1.5 (-2.9, -0.2) |
| | ↳ 2100 | | 3.3 (-0.3, 5.9) | 1.8 (1.4, 2.4) | -1.5 (-5.7, -0.9) | 0.9 (0.3, 3) | -2.1 (-7, 0) |
| | RCP1.9 in 2050 | 4/5 | -2.2 (-7, 0.6) | 4.5 (2.1, 7) | -1.2 (-2, 0.3) | 3.4 (-0.9, 7) | -4.8 (-6.2, -0.4) |
| | ↳ 2100 | | -2.3 (-9.6, 2.7) | 6.6 (3.6, 11) | -2.9 (-4, 0.1) | 6.4 (-0.8, 9.5) | -7.6 (-11.7, -1.3) |
| | RCP2.6 in 2050 | 5/5 | -3.2 (-4.2, 0.1) | 2.2 (1.7, 4.7) | 0.6 (-1.9, 1.9) | 1.6 (-0.9, 4.2) | -1.4 (-3.7, 0.4) |
| | ↳ 2100 | | -5.2 (-7.2, 0.5) | 6.9 (2.3, 10.8) | -1.4 (-4, 0.8) | 5.6 (-0.9, 5.9) | -7.2 (-8, 0.5) |
| SSP3 | RCP4.5 in 2050 | 5/5 | -2.2 (-2.2, 0.7) | 1.5 (0.1, 2.1) | 1.2 (-0.9, 2.7) | -0.9 (-2.5, 2.9) | -0.1 (-2.5, 1.6) |
| | ↳ 2100 | | -3.4 (-4.7, 1.5) | 4.1 (0.4, 6.3) | 0.7 (-2.6, 3.1) | -0.5 (-3.1, 5.9) | -2.8 (-5.3, 1.9) |
| | Baseline in 2050 | 5/5 | -1.5 (-2.6, -0.2) | 0.7 (0, 1.5) | 1.3 (1, 2.7) | -1.3 (-2.5, -0.4) | -0.1 (-1.2, 1.6) |
| | ↳ 2100 | | -2.1 (-5.9, 0.3) | 1.2 (0.1, 2.4) | 1.9 (0.8, 2.8) | -1.3 (-2.7, -0.2) | -0.2 (-1.9, 2.1) |
| | RCP1.9 in 2050 | Infeasible in all assessed models | | - | - | - | - |
| | ↳ 2100 | | | - | - | - | - |
| SSP4 | RCP2.6 in 2050 | Infeasible in all assessed models | | - | - | - | - |
| | ↳ 2100 | | | - | - | - | - |
| | RCP4.5 in 2050 | 3/3 | -3.4 (-4.4, -2) | 1.3 (1.3, 2) | 2.3 (1.2, 3) | -2.4 (-4, -1) | 2.1 (-0.1, 3.8) |
| | ↳ 2100 | | -6.2 (-6.8, -5.4) | 4.6 (1.5, 7.1) | 3.4 (1.9, 4.5) | -3.1 (-5.5, -0.3) | 2 (-2.5, 4.4) |
| | Baseline in 2050 | 4/4 | -3 (-4.6, -1.7) | 1 (0.2, 1.5) | 2.5 (1.5, 3) | -2.5 (-4, -1.5) | 2.4 (0.6, 3.8) |
| | ↳ 2100 | | -5 (-7.1, -4.2) | 1.1 (0.9, 2.5) | 5.1 (3.8, 6.1) | -5.3 (-6, -2.6) | 3.4 (0.9, 6.4) |
| SSP5 | RCP1.9 in 2050 | Infeasible in all assessed models** | | - | - | - | - |
| | ↳ 2100 | | | - | - | - | - |
| | RCP2.6 in 2050 | 3/3 | -4.5 (-6, -2.1) | 3.3 (1.5, 4.5) | 0.5 (-0.1, 0.9) | 0.7 (-0.3, 2.2) | -0.6 (-0.7, 0.1) |
| | ↳ 2100 | | -5.8 (-10.2, -4.7) | 2.5 (2.3, 15.2) | -0.8 (-0.8, 1.8) | 1.4 (-1.7, 4.1) | -1.2 (-2.5, -0.2) |
| | RCP4.5 in 2050 | 3/3 | -2.7 (-4.4, -0.4) | 1.7 (1, 1.9) | 1.1 (-0.1, 1.7) | -1.8 (-2.3, 2.1) | 0.8 (-0.5, 1.5) |
| | ↳ 2100 | | -2.8 (-7.8, -2) | 2.7 (2.3, 4.7) | 1.1 (0.2, 1.2) | -0.7 (-2.6, 1) | 1.4 (-1, 1.8) |
| SSP5 | Baseline in 2050 | 3/3 | -2.8 (-2.9, -0.2) | 1.1 (0.7, 2) | 1.1 (0.7, 1.8) | -1.8 (-2.3, -1) | 1.5 (-0.5, 2.1) |
| | ↳ 2100 | | -2.4 (-5, -1) | 1.7 (1.4, 2.6) | 1.2 (1.2, 1.9) | -2.4 (-2.5, -2) | 1.3 (-1, 4.4) |
| | RCP1.9 in 2050 | 2/4 | -1.5 (-3.9, 0.9) | 6.7 (6.2, 7.2) | -1.9 (-3.5, -0.4) | 3.1 (-0.1, 6.3) | -6.4 (-7.7, -5.1) |
| | ↳ 2100 | | -0.5 (-4.2, 3.2) | 7.6 (7.2, 8) | -3.4 (-6.2, -0.5) | 4.7 (0.1, 9.4) | -8.5 (-10.7, -6.2) |
| | RCP2.6 in 2050 | 4/4 | -3.4 (-6.9, 0.3) | 4.8 (3.8, 5.1) | -2.1 (-4, 1) | 3.9 (-0.1, 6.7) | -4.4 (-5, 0.2) |
| | ↳ 2100 | | -4.3 (-8.4, 0.5) | 9.1 (7.7, 9.2) | -3.3 (-6.5, -0.5) | 3.9 (-0.1, 9.3) | -6.3 (-9.1, -1.4) |
| SSP5 | RCP4.5 in 2050 | 4/4 | -2.5 (-3.7, 0.2) | 1.7 (0.6, 2.9) | 0.6 (-3.3, 1.9) | -0.1 (-1.7, 6) | -1.2 (-2.6, 2.3) |
| | ↳ 2100 | | -4.1 (-4.6, 0.7) | 4.8 (2, 8) | -1 (-5.5, 1) | -0.2 (-1.4, 9.1) | -3 (-5.2, 2.1) |
| | Baseline in 2050 | 4/4 | -0.6 (-3.8, 0.4) | 0.8 (0, 2.1) | 1.5 (-0.7, 3.3) | -1.9 (-3.4, 0.5) | -0.1 (-1.5, 2.9) |
| | ↳ 2100 | | -0.2 (-2.4, 1.8) | 1 (0.2, 2.3) | 1 (-2, 2.5) | -2.1 (-3.4, 1.1) | -0.4 (-2.4, 2.8) |

* Count of models included / Count of models attempted. One model did not provide land data and is excluded from all entries.

** One model could reach RCP1.9 with SSP4, but did not provide land data

1 **Chapter 1: Framing and Context**

2

3 **Coordinating Lead Authors:** Almut Arneth (Germany), Fatima Denton (The Gambia)

4 **Lead Authors:** Fahmuddin Agus (Indonesia), Aziz Elbehri (Morocco), Karlheinz Erb (Italy), Balgis
5 Osman Elasha (Côte d'Ivoire), Mohammad Rahimi (Iran), Mark Rounsevell (United Kingdom), Adrian
6 Spence (Jamaica), Riccardo Valentini (Italy)

7 **Contributing Authors:** Peter Alexander (United Kingdom), Yuping Bai (China), Ana Bastos
8 (Portugal/Germany), Niels Debonne (The Netherlands), Jan Fuglestvedt (Norway), Rafaela Hillerbrand
9 (Germany), Baldur Janz (Germany), Thomas Kastner (Austria), Ylva Longva (United Kingdom), Patrick
10 Meyfroidt (Belgium), Michael O'Sullivan (United Kingdom)

11 **Review Editors:** Edvin Aldrian (Indonesia), Bruce McCarl (The United States of America), Maria Jose
12 Sanz Sanchez (Spain)

13 **Chapter Scientists:** Yuping Bai (China), Baldur Janz (Germany)

14 **Date of Draft:** 07/08/2019

15

| | | |
|----|--|------|
| 1 | Table of Contents | |
| 2 | | |
| 3 | Chapter 1: Framing and Context..... | 1-1 |
| 4 | Executive summary..... | 1-1 |
| 5 | 1.1 Introduction and scope of the report | 1-3 |
| 6 | 1.1.1 Objectives and scope of the assessment..... | 1-3 |
| 7 | 1.1.2 Status and dynamics of the (global) land system | 1-7 |
| 8 | 1.2 Key challenges related to land use change..... | 1-13 |
| 9 | 1.2.1 Land system change, land degradation, desertification and food security | 1-13 |
| 10 | 1.2.2 Progress in dealing with uncertainties in assessing land processes in the climate system... 1- | |
| 11 | 18 | |
| 12 | Cross-Chapter Box 1: Scenarios and other methods to characterise the future of land | 1-22 |
| 13 | 1.3 Response options to the key challenges | 1-26 |
| 14 | 1.3.1 Targeted decarbonisation relying on large land-area need..... | 1-27 |
| 15 | Cross-Chapter Box 2: Implications of large-scale conversion from non-forest to forest land..... | 1-29 |
| 16 | 1.3.2 Land Management..... | 1-32 |
| 17 | 1.3.3 Value chain management | 1-32 |
| 18 | 1.3.4 Risk management | 1-35 |
| 19 | 1.3.5 Economics of land-based mitigation pathways: Costs versus benefits of early action under | |
| 20 | uncertainty..... | 1-35 |
| 21 | 1.3.6 Adaptation measures and scope for co-benefits with mitigation | 1-36 |
| 22 | 1.4 Enabling the response | 1-37 |
| 23 | 1.4.1 Governance to enable the response..... | 1-37 |
| 24 | 1.4.2 Gender agency as a critical factor in climate and land sustainability outcomes | 1-39 |
| 25 | 1.4.3 Policy Instruments..... | 1-40 |
| 26 | 1.5 The interdisciplinary nature of the SRCCL | 1-42 |
| 27 | Frequently Asked Questions | 1-42 |
| 28 | References..... | 1-44 |
| 29 | Supplementary Material..... | 1-90 |
| 30 | | |
| 31 | | |

1 Executive summary

2 **Land, including its water bodies, provides the basis for human livelihoods and well-being through**
3 **primary productivity, the supply of food, freshwater, and multiple other ecosystem services (*high***
4 ***confidence*).** Neither our individual or societal identities, nor the World's economy would exist without
5 the multiple resources, services and livelihood systems provided by land ecosystems and biodiversity.
6 The annual value of the World's total terrestrial ecosystem services has been estimated at 75–85 trillion
7 USD in 2011 (based on USD 2007 values) (*low confidence*). This substantially exceeds the annual World
8 GDP (*high confidence*). Land and its biodiversity also represent essential, intangible benefits to humans,
9 such as cognitive and spiritual enrichment, sense of belonging and aesthetic and recreational values.
10 Valuing ecosystem services with monetary methods often overlooks these intangible services that shape
11 societies, cultures and quality of life and the intrinsic value of biodiversity. The Earth's land area is finite.
12 Using land resources sustainably is fundamental for human well-being (*high confidence*). {1.1.1}

13 **The current geographic spread of the use of land, the large appropriation of multiple ecosystem**
14 **services and the loss of biodiversity are unprecedented in human history (*high confidence*).** By 2015,
15 about three-quarters of the global ice-free land surface was affected by human use. Humans appropriate
16 one quarter to one third of global terrestrial potential net primary production (*high confidence*). Croplands
17 cover 12–14% of the global ice-free surface. Since 1961, the supply of global per capita food calories
18 increased by about one third, with the consumption of vegetable oils and meat more than doubling. At the
19 same time, the use of inorganic nitrogen fertiliser increased by nearly 9-fold, and the use of irrigation
20 water roughly doubled (*high confidence*). Human use, at varying intensities, affects about 60–85% of
21 forests and 70–90% of other natural ecosystems (e.g., savannahs, natural grasslands) (*high confidence*).
22 Land use caused global biodiversity to decrease by around 11–14% (*medium confidence*). {1.1.2}

23 **Warming over land has occurred at a faster rate than the global mean and this has had observable**
24 **impacts on the land system (*high confidence*).** The average temperature over land for the period 1999–
25 2018 was 1.41°C higher than for the period 1881–1900, and 0.54°C larger than the equivalent global
26 mean temperature change. These warmer temperatures (with changing precipitation patterns) have altered
27 the start and end of growing seasons, contributed to regional crop yield reductions, reduced freshwater
28 availability, and put biodiversity under further stress and increased tree mortality (*high confidence*).
29 Increasing levels of atmospheric CO₂, have contributed to observed increases in plant growth as well as to
30 increases in woody plant cover in grasslands and savannahs (*medium confidence*). {1.1.2}

31 **Urgent action to stop and reverse the over-exploitation of land resources would buffer the negative**
32 **impacts of multiple pressures, including climate change, on ecosystems and society (*high***
33 ***confidence*).** Socio-economic drivers of land use change such as technological development, population
34 growth and increasing per capita demand for multiple ecosystem services are projected to continue into
35 the future (*high confidence*). These and other drivers can amplify existing environmental and societal
36 challenges, such as the conversion of natural ecosystems into managed land, rapid urbanisation, pollution
37 from the intensification of land management and equitable access to land resources (*high confidence*).
38 Climate change will add to these challenges through direct, negative impacts on ecosystems and the
39 services they provide (*high confidence*). Acting immediately and simultaneously on these multiple drivers
40 would enhance food, fibre and water security, alleviate desertification, and reverse land degradation,
41 without compromising the non-material or regulating benefits from land (*high confidence*). {1.1.2, 1.2.1,
42 1.3.2-1.3.6, Cross-Chapter Box 1, Chapter 1}

1 **Rapid reductions in anthropogenic greenhouse gas emissions that restrict warming to “well-below”**
2 **2°C would greatly reduce the negative impacts of climate change on land ecosystems (*high***
3 ***confidence*).** In the absence of rapid emissions reductions, reliance on large-scale, land-based,
4 **climate change mitigation is projected to increase, which would aggravate existing pressures on**
5 **land (*high confidence*).** Climate change mitigation efforts that require large land areas (e.g., bioenergy
6 and afforestation/reforestation) are projected to compete with existing uses of land (*high confidence*). The
7 competition for land could increase food prices and lead to further intensification (e.g., fertiliser and water
8 use) with implications for water and air pollution, and the further loss of biodiversity (*medium*
9 *confidence*). Such consequences would jeopardise societies’ capacity to achieve many sustainable
10 development goals that depend on land (*high confidence*). {1.3.1, Cross-Chapter Box 2 in Chapter 1}

11 **Nonetheless, there are many land-related climate change mitigation options that do not increase the**
12 **competition for land (*high confidence*).** Many of these options have co-benefits for climate change
13 **adaptation (*medium confidence*).** Land use contributes about one quarter of global greenhouse gas
14 emissions, notably CO₂ emissions from deforestation, CH₄ emissions from rice and ruminant livestock
15 and N₂O emissions from fertiliser use (*high confidence*). Land ecosystems also take up large amounts of
16 carbon (*high confidence*). Many land management options exist to both reduce the magnitude of
17 emissions and enhance carbon uptake. These options enhance crop productivity, soil nutrient status,
18 microclimate or biodiversity, and thus, support adaptation to climate change (*high confidence*). In
19 addition, changes in consumer behaviour, such as reducing the over-consumption of food and energy
20 would benefit the reduction of GHG emissions from land (*high confidence*). The barriers to the
21 implementation of mitigation and adaptation options include skills deficit, financial and institutional
22 barriers, absence of incentives, access to relevant technologies, consumer awareness and the limited
23 spatial scale at which the success of these practices and methods have been demonstrated. {1.2.1, 1.3.2,
24 1.3.3, 1.3.4, 1.3.5, 1.3.6}

25 **Sustainable food supply and food consumption, based on nutritionally balanced and diverse diets,**
26 **would enhance food security under climate and socio-economic changes (*high confidence*).**
27 Improving food access, utilisation, quality and safety to enhance nutrition, and promoting globally
28 equitable diets compatible with lower emissions have demonstrable positive impacts on land use and food
29 security (*high confidence*). Food security is also negatively affected by food loss and waste (estimated as
30 more than 30% of harvested materials) (*high confidence*). Barriers to improved food security include
31 economic drivers (prices, availability and stability of supply) and traditional, social and cultural norms
32 around food eating practices. Climate change is expected to increase variability in food production and
33 prices globally (*high confidence*), but the trade in food commodities can buffer these effects. Trade can
34 provide embodied flows of water, land and nutrients (*medium confidence*). Food trade can also have
35 negative environmental impacts by displacing the effects of overconsumption (*medium confidence*).
36 Future food systems and trade patterns will be shaped as much by policies as by economics (*medium*
37 *confidence*). {1.2.1, 1.3.3}

38 **A gender inclusive approach offers opportunities to enhance the sustainable management of land**
39 **(*medium confidence*).** Women play a significant role in agriculture and rural economies globally. In
40 many World regions, laws, cultural restrictions, patriarchy and social structures such as discriminatory
41 customary laws and norms reduce women’s capacity in supporting the sustainable use of land resources
42 (*medium confidence*). Therefore, acknowledging women’s land rights and bringing women’s land
43 management knowledge into land-related decision-making would support the alleviation of land
44 degradation, and facilitate the take-up of integrated adaptation and mitigation measures (*medium*
45 *confidence*). {1.4.1, 1.4.2}

1 **Regional and country specific contexts affect the capacity to respond to climate change and its**
2 **impacts, through adaptation and mitigation (*high confidence*).** There is large variability in the
3 availability and use of land resources between regions, countries and land-management systems. In
4 addition, differences in socio-economic conditions, such as wealth, degree of industrialisation, institutions
5 and governance, affect the capacity to respond to climate change, food insecurity, land degradation and
6 desertification. The capacity to respond is also strongly affected by local land ownership. Hence, climate
7 change will affect regions and communities differently (*high confidence*). {1.3, 1.4}

8 **Cross-scale, cross-sectoral and inclusive governance can enable coordinated policy that supports**
9 **effective adaptation and mitigation (*high confidence*).** There is a lack of coordination across
10 governance levels, for example, local, national, transboundary and international, in addressing climate
11 change and sustainable land management challenges. Policy design and formulation is often strongly
12 sectoral, which poses further barriers when integrating international decisions into relevant (sub)national
13 policies. A portfolio of policy instruments that are inclusive of the diversity of governance actors would
14 enable responses to complex land and climate challenges (*high confidence*). Inclusive governance that
15 considers women's and indigenous people's rights to access and use land enhances the equitable sharing
16 of land resources, fosters food security and increases the existing knowledge about land use, which can
17 increase opportunities for adaptation and mitigation (*medium confidence*). {1.3.5, 1.4.1, 1.4.2, 1.4.3}

18 **Scenarios and models are important tools to explore the trade-offs and co-benefits of land**
19 **management decisions under uncertain futures (*high confidence*).** Participatory, co-creation processes
20 with stakeholders can facilitate the use of scenarios in designing future sustainable development strategies
21 (*medium confidence*). In addition to qualitative approaches, models are critical in quantifying scenarios,
22 but uncertainties in models arise from, for example, differences in baseline datasets, land cover classes
23 and modelling paradigms (*medium confidence*). Current scenario approaches are limited in quantifying
24 time-dependent, policy and management decisions that can lead from today to desirable futures or visions.
25 Advances in scenario analysis and modelling are needed to better account for full environmental costs and
26 non-monetary values as part of human decision-making processes. {1.2.2, Cross Chapter Box 1 in
27 Chapter 1}

28 **1.1 Introduction and scope of the report**

29 **1.1.1 Objectives and scope of the assessment**

30 Land, including its water bodies, provides the basis for our livelihoods through basic processes such as
31 net primary production that fundamentally sustain the supply of food, bioenergy and freshwater, and the
32 delivery of multiple other ecosystem services and biodiversity (Hoekstra and Wiedmann 2014; Mace et
33 al. 2012; Newbold et al. 2015; Runting et al. 2017; Isbell et al. 2017)(see Cross-Chapter Box 8:
34 Ecosystem Services, Chapter 6). The annual value of the world's total terrestrial ecosystem services has
35 been estimated to be about USD 75–85 trillion (in 2011 based on USD 2007 values)(Costanza et al.
36 2014). This equates approximately to the world's average GDP over the last 5 years (IMF 2018). Land
37 also supports non-material ecosystem services such as cognitive and spiritual enrichment and aesthetic
38 values (Hernández-Morcillo et al. 2013; Fish et al. 2016), intangible services that shape societies, cultures
39 and human well-being. Exposure of people living in cities to (semi-)natural environments has been found
40 to decrease mortality, cardiovascular disease and depression (Rook 2013; Terraube et al. 2017). Non-
41 material and regulating ecosystem services have been found to decline globally and rapidly, often at the
42 expense of increasing material services (Fischer et al. 2018; IPBES 2018a). Climate change will
43 exacerbate diminishing land and freshwater resources, increase biodiversity loss, and will intensify

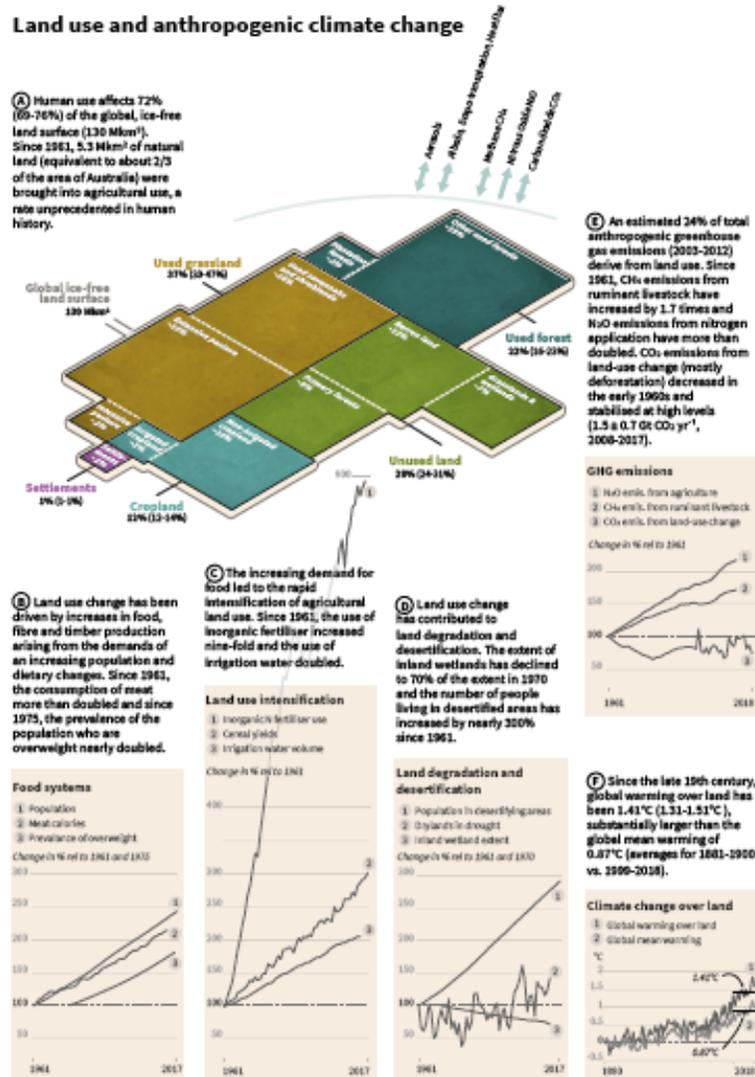
1 societal vulnerabilities, especially in regions where economies are highly dependent on natural resources.
2 Enhancing food security and reducing malnutrition, whilst also halting and reversing desertification and
3 land degradation, are fundamental societal challenges that are increasingly aggravated by the need to both
4 adapt to and mitigate climate change impacts without compromising the non-material benefits of land
5 (Kongsager et al. 2016; FAO et al. 2018).

6 Annual emissions of greenhouse gases (GHGs) and other climate forcers continue to increase unabatedly.
7 *Confidence is very high* that the window of opportunity, the period when significant change can be made,
8 for limiting climate change within tolerable boundaries is rapidly narrowing (Schaeffer et al. 2015;
9 Bertram et al. 2015; Riahi et al. 2015; Millar et al. 2017; Rogelj et al. 2018a). The Paris Agreement
10 formulates the goal of limiting global warming this century well below 2°C above pre-industrial levels,
11 for which rapid actions are required across the energy, transport, infrastructure and agricultural sectors,
12 while factoring in the need for these sectors to accommodate a growing human population (Wynes and
13 Nicholas 2017; Le Quere et al. 2018). Conversion of natural land, and land management, are significant
14 net contributors to GHG emissions and climate change, but land ecosystems are also a GHG sink (Smith
15 et al. 2014; Tubiello et al. 2015; Le Quere et al. 2018; Ciais et al. 2013a). It is not surprising, therefore,
16 that land plays a prominent role in many of the Nationally Determined Contributions (NDCs) of the
17 parties to the Paris Agreement (Rogelj et al. 2018a,b; Grassi et al. 2017; Forsell et al. 2016), and land-
18 measures will be part of the NDC review by 2023.

19 A range of different climate change mitigation and adaptation options on land exist, which differ in terms
20 of their environmental and societal implications (Meyfroidt 2018; Bonsch et al. 2016; Crist et al. 2017;
21 Humpenoder et al. 2014; Harvey and Pilgrim 2011; Mouratiadou et al. 2016; Zhang et al. 2015; Sanz-
22 Sanchez et al. 2017; Pereira et al. 2010; Griscom et al. 2017; Rogelj et al. 2018a)(see Chapters 4-6). The
23 Special Report on climate change, desertification, land degradation, sustainable land management, food
24 security, and GHG fluxes in terrestrial ecosystems (SRCCL) synthesises the current state of scientific
25 knowledge on the issues specified in the report's title (see Figure 1.1, Figure 1.2). This knowledge is
26 assessed in the context of the Paris Agreement, but many of the SRCCL issues concern other international
27 conventions such as the United Nations Convention on Biodiversity (UNCBD), the UN Convention to
28 Combat Desertification (UNCCD), the UN Sendai Framework for Disaster Risk Reduction (UNISDR)
29 and the UN Agenda 2030 and its Sustainable Development Goals (SDGs). The SRCCL is the first report
30 in which land is the central focus since the IPCC Special Report on land use, land-use change and forestry
31 (Watson et al. 2000)(see Box 1.1). The main objectives of the SRCCL are to:

- 32 1) Assess the current state of the scientific knowledge on the impacts of socio-economic drivers and their
33 interactions with climate change on land, including degradation, desertification and food security;
- 34 2) Evaluate the feasibility of different land-based response options to GHG mitigation, and assess the
35 potential synergies and trade-offs with ecosystem services and sustainable development;
- 36 3) Examine adaptation options under a changing climate to tackle land degradation and desertification
37 and to build resilient food systems, as well as evaluating the synergies and trade-offs between
38 mitigation and adaptation; and
- 39 4) Delineate the policy, governance and other enabling conditions to support climate mitigation, land
40 ecosystem resilience and food security in the context of risks, uncertainties and remaining knowledge
41 gaps.

42

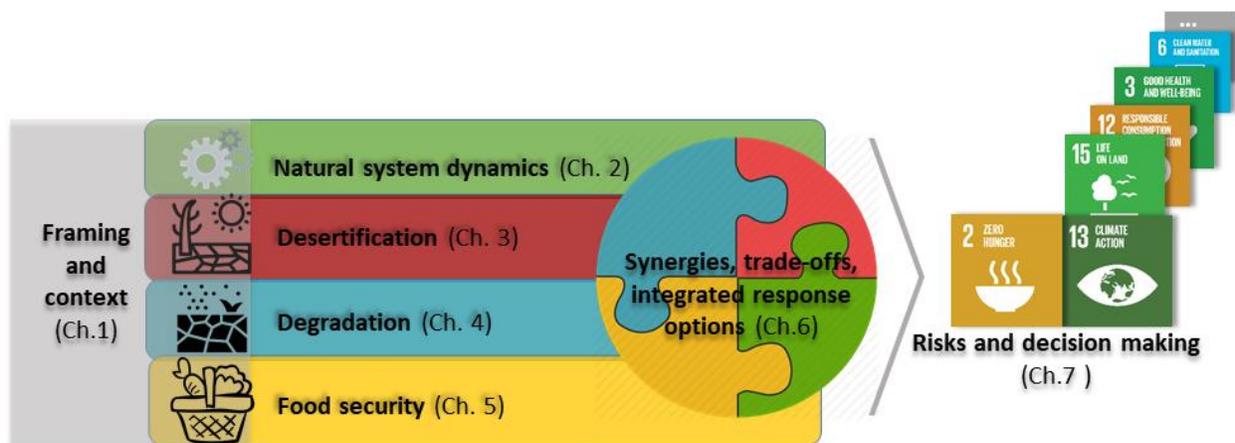


1

2 **Figure 1.1 A representation of the principal land challenges and land-climate system processes covered in this**
 3 **assessment report.** A. The tiles show the current extent (in about 2015) of the human use of the land surface,
 4 aggregated into five broad land use and land cover categories with uncertainty ranges. Colour shading indicates
 5 different intensities of human use (Table 1.1). B. Agricultural areas have increased to supply the increasing demand
 6 for food arising from population growth, income growth and increasing consumption of animal-sourced products.
 7 The proportion of the global population that is overweight (body mass index > 25 kg/m²) has increased markedly
 8 (section 5.1.2). Population density (*Source: United Nations, Department of Economic and Social Affairs 2017*).
 9 Meat calories supplied (*Source: FAOSTAT 2018*) Prevalence of people overweight (*Source: Abarca-Gómez et al.*
 10 *2017*)(5.1.2). C. Increasing food production has led to rapid land use intensification, including increases in the use of
 11 nitrogen fertiliser and irrigation water that have supported the growth in cereal yields (section 1.1). Change in cereal
 12 yield and irrigation water use (*Source: FAOSTAT 2018*); Change in total inorganic nitrogen fertiliser consumption
 13 (*Source: International Fertiliser Industry Association, <https://www.ifastat.org/databases>*). Note that the very large
 14 percentage change in fertiliser use reflects the very low use in 1961. The increase relates to both increasing fertiliser
 15 input per area as well as the expansion of fertilised cropland and grassland. D. Land use change has led to
 16 substantial losses in the extent of inland wetlands (section 4.2.1, 4.6.1). Dryland areas are under increasing pressures
 17 both from the increasing number of people living in these areas and from the increase in droughts (section 3.1.1).
 18 The inland wetland extent trends (WET) index was developed by aggregating data from 2130 time series that report
 19 changes in local wetland area over time (Dixon et al. 2016; Darrah et al. 2019). Dryland areas were defined using
 20 TerraClimate precipitation and potential evapotranspiration (1980–2015) (Abatzoglou et al. 2018) to identify areas

1 where the Aridity Index is below 0.65. Areas undergoing human caused desertification, after accounting for
 2 precipitation variability and CO₂ fertilisation, are identified in (Le et al. 2016). Population data for these areas were
 3 extracted from the gridded historical population database HYDE3.2 (Goldewijk et al. 2017). The 12-month
 4 accumulation Global Precipitation Climatology Centre Drought Index (Ziese et al. 2014) was extracted for drylands.
 5 The area in drought was calculated for each month (Drought Index below -1), and the mean over the year was used
 6 to calculate the percentage of drylands in drought that year. E. Land use change and intensification have contributed
 7 to CH₄ emissions from ruminant livestock, agricultural N₂O emissions and CO₂ emissions from net deforestation
 8 {2.3}. Sources: N₂O from agricultural activities and CH₄ from enteric fermentation: Edgar database
 9 (<http://edgar.jrc.ec.europa.eu/overview.php?v=42FT2012>) from 1970. From 1970 back to 1961, CH₄ and N₂O were
 10 extrapolated using a regression with time, taken for the years 1970-1979 from Edgar. Net-land use change emissions
 11 of CO₂ are from the annual Global Carbon Budget, using the mean of two bookkeeping models (Le Quéré et al.
 12 2018). Chapter 2 (Section 2.2, 2.3) and Chapter 5 (Section 5.4) provides a discussion of uncertainties and other
 13 emissions estimates. The various exchanges between the land surface and the atmosphere, including the emission
 14 and uptake of greenhouse gases, exchanges related to the land-surface energy balance and aerosols are indicated by
 15 arrows (section 2.1, 2.3, 2.4). Warming over land is more rapid than the global mean temperature change (section
 16 2.2). Future climate change will exacerbate the already considerable challenges faced by land systems. The warming
 17 curves are averages of four historical estimates, and described in Section 2.1.

18 The SRCLL identifies and assesses land-related challenges and response-options in an integrative way,
 19 aiming to be policy relevant across sectors. Chapter 1 provides a synopsis of the main issues addressed in
 20 this report, which are explored in more detail in Chapters 2–7. Chapter 1 also introduces important
 21 concepts and definitions and highlights discrepancies with previous reports that arise from different
 22 objectives (a full set of definitions is provided in the Glossary). Chapter 2 focuses on the natural system
 23 dynamics, assessing recent progress towards understanding the impacts of climate change on land, and the
 24 feedbacks arising from altered biogeochemical and biophysical exchange fluxes (Figure 1.2).



25

26

Figure 1.2 Overview over the SRCLL

27 Chapter 3 examines how the world's dryland populations are uniquely vulnerable to desertification and
 28 climate change, but also have significant knowledge in adapting to climate variability and addressing
 29 desertification. Chapter 4 assesses the urgency of tackling land degradation across all land ecosystems.
 30 Despite accelerating trends of land degradation, reversing these trends is attainable through restoration
 31 efforts and proper implementation of sustainable land management (SLM), which is expected to improve
 32 resilience to climate change, mitigate climate change, and ensure food security for generations to come.
 33 Food security is the focus of Chapter 5, with an assessment of the risks and opportunities that climate
 34 change presents to food systems, considering how mitigation and adaptation can contribute to both human
 35 and planetary health.

1 Chapters 6 focuses on the response options within the land system that deal with trade-offs and increase
2 benefits in an integrated way in support of the SDGs. Chapter 7 highlights these aspects further, by
3 assessing the opportunities, decision making and policy responses to risks in the climate-land-human
4 system.

5

6 **Box 1.1 Land in previous IPCC and other relevant reports**

7 Previous IPCC reports have made reference to land and its role in the climate system. Threats to
8 agriculture forestry and other ecosystems, but also the role of land and forest management in climate
9 change, have been documented since the IPCC Second Assessment Report, especially so in the Special
10 report on land use, land-use change and forestry (Watson et al. 2000). The IPCC Special Report on
11 Extreme events (SREX) discussed sustainable land management, including land use planning, and
12 ecosystem management and restoration among the potential low-regret measures that provide benefits
13 under current climate and a range of future, climate change scenarios. Low-regret measures are defined in
14 the report as those with the potential to offer benefits now and lay the foundation for tackling future,
15 projected change. Compared to previous IPCC reports, the SRCCL offers a more integrated analysis of
16 the land system as it embraces multiple direct and indirect drivers of natural resource management
17 (related to food, water and energy securities), which have not previously been addressed to a similar depth
18 (Field et al. 2014a; Edenhofer et al. 2014).

19 The recent IPCC Special Report on Global Warming of 1.5°C (SR15) targeted specifically the Paris
20 Agreement, without exploring the possibility of future global warming trajectories above 2°C (IPCC
21 2018). Limiting global warming to 1.5°C compared to 2°C is projected to lower the impacts on terrestrial,
22 freshwater and coastal ecosystems and to retain more of their services for people. In many scenarios
23 proposed in this report, large-scale land use features as a mitigation measure. In the reports of the Food
24 and Agriculture Organisation (FAO), land degradation is discussed in relation to ecosystem goods and
25 services, principally from a food security perspective (FAO and ITPS 2015). The UNCCD report (2014)
26 discusses land degradation through the prism of desertification. It devotes due attention to how land
27 management can contribute to reversing the negative impacts of desertification and land degradation. The
28 IPBES assessments (2018a,b,c,d,e) focuses on biodiversity drivers, including a focus on land degradation
29 and desertification, with poverty as a limiting factor. The reports draw attention to a world in peril in
30 which resource scarcity conspires with drivers of biophysical and social vulnerability to derail the
31 attainment of sustainable development goals. As discussed in chapter 4 of the SRCCL, different
32 definitions of degradation have been applied in the IPBES degradation assessment (IPBES 2018b), which
33 potentially can lead to different conclusions for restoration and ecosystem management.

34 The SRCCL complements and adds to previous assessments, whilst keeping the IPCC-specific “climate
35 perspective”. It includes a focussed assessment of risks arising from maladaptation and land-based
36 mitigation (i.e. not only restricted to direct risks from climate change impacts) and the co-benefits and
37 trade-offs with sustainable development objectives. As the SRCCL cuts across different policy sectors it
38 provides the opportunity to address a number of challenges in an integrative way at the same time, and it
39 progresses beyond other IPCC reports in having a much more comprehensive perspective on land.

40 **1.1.2 Status and dynamics of the (global) land system**

41 **1.1.2.1 Land ecosystems and climate change**

42 Land ecosystems play a key role in the climate system, due to their large carbon pools and carbon
43 exchange fluxes with the atmosphere (Ciais et al. 2013b). Land use, the total of arrangements, activities

1 and inputs applied to a parcel of land (such as agriculture, grazing, timber extraction, conservation or city
2 dwelling; see glossary), and land management (sum of land-use practices that take place within broader
3 land-use categories, see glossary) considerably alter terrestrial ecosystems and play a key role in the
4 global climate system. An estimated one quarter of total anthropogenic GHG emissions arise mainly from
5 deforestation, ruminant livestock and fertiliser application (Smith et al. 2014; Tubiello et al. 2015; Le
6 Quere et al. 2018; Ciais et al. 2013a), and especially methane and nitrous oxide emissions from
7 agriculture have been rapidly increasing over the last decades (Hoesly et al. 2018; Tian et al. 2019)(see
8 Figure 1.1, see Section 2.3.2, 2.3.3).

9 Globally, land also serves as a large carbon dioxide sink, which was estimated for the period 2008–2017
10 to be nearly 30% of total anthropogenic emissions (Le Quere et al. 2015; Canadell and Schulze 2014;
11 Ciais et al. 2013a; Zhu et al. 2016)(see Section 2.3.1). This sink has been attributed to increasing
12 atmospheric CO₂ concentration, a prolonged growing season in cool environments, or forest regrowth (Le
13 Quéré et al. 2013; Pugh et al. 2019; Le Quéré et al. 2018; Ciais et al. 2013a; Zhu et al. 2016). Whether or
14 not this sink will persist into the future is one of the largest uncertainties in carbon cycle and climate
15 modelling (Ciais et al. 2013a; Bloom et al. 2016; Friend et al. 2014; Le Quere et al. 2018). In addition,
16 changes in vegetation cover caused by land use (such as conversion of forest to cropland or grassland, and
17 vice versa) can result in regional cooling or warming through altered energy and momentum transfer
18 between ecosystems and the atmosphere. Regional impacts can be substantial, but whether the effect leads
19 to warming or cooling depends on the local context (Lee et al. 2011; Zhang et al. 2014; Alkama and
20 Cescatti 2016; see Section 2.6). Due to the current magnitude of GHG emissions and carbon dioxide
21 removal in land ecosystems, there is *high confidence* that greenhouse-gas reduction measures in
22 agriculture, livestock management and forestry would have substantial climate change mitigation
23 potential with co-benefits for biodiversity and ecosystem services (Smith and Gregory 2013; Smith et al.
24 2014; Griscom et al. 2017; see Section 2.6, Section 6.3).

25 The mean temperature increase over land has been substantially larger than the global mean (land and
26 ocean), averaging 1.41°C vs. 0.87°C for the years 1999–2018 compared with 1881–1900 (see Section
27 2.2). Climate change affects land ecosystems in various ways (see Section 7.2). Growing seasons and
28 natural biome boundaries shift in response to warming or changes in precipitation (Gonzalez et al. 2010;
29 Wärlind et al. 2014; Davies-Barnard et al. 2015; Nakamura et al. 2017). Atmospheric CO₂ increases have
30 been attributed to underlie, at least partially, observed woody plant cover increase in grasslands and
31 savannahs (Donohue et al. 2013). Climate change-induced shifts in habitats, together with warmer
32 temperatures, causes pressure on plants and animals (Pimm et al. 2014; Urban et al. 2016). National
33 cereal crop losses of nearly 10% have been estimated for the period 1964–2007 as a consequence of heat
34 and drought weather extremes (Deryng et al. 2014; Lesk et al. 2016). Climate change is expected to
35 reduce yields in areas that are already under heat and water stress (Schlenker and Lobell 2010; Lobell et
36 al. 2011,2012; Challinor et al. 2014; see Section 5.2.2). At the same time, warmer temperatures can
37 increase productivity in cooler regions (Moore and Lobell 2015) and might open opportunities for crop
38 area expansion, but any overall benefits might be counterbalanced by reduced suitability in warmer
39 regions (Pugh et al. 2016; Di Paola et al. 2018). Increasing atmospheric CO₂ is expected to increase
40 productivity and water use efficiency in crops and in forests (Muller et al. 2015; Nakamura et al. 2017;
41 Kimball 2016). The increasing number of extreme weather events linked to climate change is also
42 expected to result in forest losses; heat waves and droughts foster wildfires (Seidl et al. 2017; Fasullo et
43 al. 2018; see Cross-Chapter Box 3: Fire and Climate Change, Chapter 2). Episodes of observed enhanced
44 tree mortality across many world regions have been attributed to heat and drought stress (Allen et al.
45 2010; Anderegg et al. 2012), whilst weather extremes also impact local infrastructure and hence

1 transportation and trade in land-related goods (Schweikert et al. 2014; Chappin and van der Lei 2014).
2 Thus, adaptation is a key challenge to reduce adverse impacts on land systems (see Section 1.3.6).

3 *1.1.2.2 Current patterns of land use and land cover*

4 Around three quarters of the global ice-free land, and most of the highly-productive land area, are by now
5 under some form of land use (Erb et al. 2016a; Luysaert et al. 2014; Venter et al. 2016; see Table 1.1).
6 One third of used land is associated with changed land cover. Grazing land is the single largest land-use
7 category, followed by used forestland and cropland. The total land area used to raise livestock is notable:
8 it includes all grazing land and an estimated additional one fifth of cropland for feed production (Foley et
9 al. 2011). Globally, 60–85% of the total forested area is used, at different levels of intensity, but
10 information on management practices globally are scarce (Erb et al. 2016a). Large areas of unused
11 (primary) forests remain only in the tropics and northern boreal zones (Luysaert et al. 2014; Birdsey and
12 Pan 2015; Morales-Hidalgo et al. 2015; Potapov et al. 2017; Erb et al. 2017), while 73–89% of other,
13 non-forested natural ecosystems (natural grasslands, savannas, etc.) are used. Large uncertainties relate to
14 the extent of forest (32.0–42.5 million km²) and grazing land (39–62 million km²), due to discrepancies in
15 definitions and observation methods (Luysaert et al. 2014; Erb et al. 2017; Putz and Redford 2010;
16 Schepaschenko et al. 2015; Birdsey and Pan 2015; FAO 2015a; Chazdon et al. 2016a; FAO 2018a).
17 Infrastructure areas (including settlements, transportation and mining), while being almost negligible in
18 terms of extent, represent particularly pervasive land-use activities, with far-reaching ecological, social
19 and economic implications (Cherlet et al. 2018; Laurance et al. 2014).

20 The large imprint of humans on the land surface has led to the definition of anthromes, i.e. large-scale
21 ecological patterns created by the sustained interactions between social and ecological drivers. The
22 dynamics of these ‘anthropogenic biomes’ are key for land-use impacts as well as for the design of
23 integrated response options (Ellis and Ramankutty 2008; Ellis et al. 2010; Cherlet et al. 2018; Ellis et al.
24 2010, see Chapter 6).

25 The intensity of land use varies hugely within and among different land use types and regions. Averaged
26 globally, around 10% of the ice-free land surface was estimated to be intensively managed (such as tree
27 plantations, high livestock density grazing, large agricultural inputs), two thirds moderately and the
28 remainder at low intensities (Erb et al. 2016a). Practically all cropland is fertilised, with large regional
29 variations. Irrigation is responsible for 70% of ground- or surface-water withdrawals by humans (Wisser
30 et al. 2008; Chaturvedi et al. 2015; Siebert et al. 2015; FAOSTAT 2018). Humans appropriate one quarter
31 to one third of the total potential net primary production, i.e. the NPP that would prevail in the absence of
32 land use (estimated at about 60 GtC yr⁻¹; Bajželj et al. 2014; Haberl et al. 2014), about equally through
33 biomass harvest and changes in NPP due to land management. The current total of agricultural (cropland
34 and grazing) biomass harvest is estimated at about 6 GtC yr⁻¹, around 50–60% of this is consumed by
35 livestock. Forestry harvest for timber and wood fuel amounts to about 1 GtC yr⁻¹ (Alexander et al. 2017;
36 Bodirsky and Müller 2014; Lassaletta et al. 2014, 2016; Mottet et al. 2017; Haberl et al. 2014; Smith et al.
37 2014; Bais et al. 2015; Bajželj et al. 2014)(see Cross-Chapter Box 7: Bioenergy and BECCS, Chapter 6).

38 **Table 1.1 Extent of global land use and management around the year 2015**

| | Best guess | Range | Range | Type | Ref. |
|---|-----------------------------|----------------|--------------|------------|--------------------|
| | [million km ²] | | [% of total] | | |
| Total | 130.4 | | 100% | | |
| USED LAND | 92.6 | 90.0-99.3 | 71% | 69-76% | |
| Infrastructure (Settlements, mining, etc.) | 1.4 | 1.2-1.9 | 1% | LCC | 1,2,3,4,5,6 |

| | | | | | | |
|--|-------------|------------------|------------|---------------|-----------|-----------------|
| Cropland | 15.9 | 15.9-18.8 | 12% | 12-14% | | 1,7 |
| irrigated cropland | 3.1 | | 2% | | LCC | 8 |
| non-irrigated cropland | 12.8 | 12.8-15.7 | 10% | | LCC | 8 |
| Grazing land | 48.0 | 38.8-61.9 | 37% | 30-47% | | |
| Permanent pastures | 27.1 | 22.8-32.8 | 21% | 17-25% | | 5,7,8 |
| Intensive permanent pastures* | 2.6 | | 2% | | LCC | 8,9 |
| Extensive perm. pastures, on potential forest sites** | 8.7 | | 7% | | LCC | 9 |
| Extensive perm. pastures, on natural grasslands** | 15.8 | 11.5-21.56 | 12% | 9-16% | LM | |
| Non-forested, used land, multiple uses[§] | 20.1 | 6.1-39.1 | 16% | 5-30% | LM | |
| Used forests[#] | 28.1 | 20.3-30.5 | 22% | 16-23% | | 10,11,12 |
| Planted forests | 2.9 | | 2% | | LCC | 12 |
| Managed for timber and other uses | 25.2 | 17.4-27.6 | 19% | 13-21% | LM | 12 |
| UNUSED LAND | 37.0 | 31.1-40.4 | 28% | 24-31% | | 5,11,13 |
| Unused, unforested ecosystems, including grasslands and wetlands | 9.4 | 5.9-10.4 | 7% | 5-8% | | 1,13 |
| Unused forests (intact or primary forests) | 12.0 | 11.7-12.0 | 9% | | | 11,12 |
| Other land (barren wilderness, rocks, etc.) | 15.6 | 13.5-18.0 | 12% | 10-14% | | 4,5,13,14 |
| Land-cover conversions (sum of LCC) | 31.5 | 31.3-34.9 | 24% | 24-27% | | |
| Land-use occurring within natural land-cover types (sum of LM) | 61.1 | 55.1-68.0 | 47% | 42-52% | | |

1 *>100 animals/km²

2 **<100 animals/km², residual category within permanent pastures

3 § Calculated as residual category. Contains land not classified as forests or cropland, such as savanna and tundra
4 used as rangelands, with extensive uses like seasonal, rough grazing, hunting, fuelwood collection outside forests,
5 wild products harvesting, etc.

6 # used forest calculated as total forest minus unused forests

7 **Note:** This table is based on data and approaches described in Lambin and Meyfroidt (2011,2014); Luyssaert et al.
8 (2014); Erb et al. (2016a), and references below. The target year for data is 2015, but proportions of some
9 subcategories are from 2000 (the year with still most reconciled datasets available) and their relative extent was
10 applied to some broad land use categories for 2015. Sources: Settlements (1): (Luyssaert et al. 2014); (2) (Lambin
11 and Meyfroidt 2014); (3) Global Human Settlements dataset, <https://ghsl.jrc.ec.europa.eu/>. Total infrastructure
12 including transportation (4) (Erb et al. 2007); (5) (Stadler et al. 2018); mining (6) (Cherlet et al. 2018); (7)
13 (FAOSTAT 2018); (8) proportions from (Erb et al. 2016a); (9) (Ramankutty et al. 2008) extrapolated from 2000 to
14 2010 trend for permanent pastures from (7); (9) (Erb et al. 2017); (10) (Schepaschenko et al. 2015); (11) (Potapov et
15 al. 2017); (12) (FAO 2015a); (13) (Venter et al. 2016); (14) (Ellis et al. 2010)

16 1.1.2.3 Past and ongoing trends

17 Globally, cropland area changed by +15% and the area of permanent pastures by +8% since the early
18 1960s (FAOSTAT 2018), with strong regional differences (Figure 1.3). In contrast, cropland production
19 since 1961 increased by about 3.5 times, the production of animal products by 2.5 times, and forestry by
20 1.5 times; in parallel with strong yield (production per unit area) increases (FAOSTAT 2018; Figure 1.3).
21 Per capita calorie supply increased by 17% (since 1970; Kastner et al. 2012), and diet composition
22 changed markedly, tightly associated with economic development and lifestyle: Since the early 1960s, per
23 capita dairy product consumption increased by a factor 1.2, and meat and vegetable oil consumption more
24 than doubled (FAO 2017, 2018b; Tilman and Clark 2014; Marques et al. 2019). Population and livestock
25 production represent key drivers of the global expansion of cropland for food production, only partly

1 compensated by yield increases at the global level (Alexander et al. 2015). A number of studies have
2 reported reduced growth rates or stagnation in yields in some regions in the last decades (*medium*
3 *evidence, high agreement*; Lin and Huybers 2012; Ray et al. 2012; Elbehri, Aziz, Joshua Elliott 2015; see
4 Section 5.2.2).

5 The past increases in agricultural production have been associated with strong increases in agricultural
6 inputs (Foley et al. 2011; Siebert et al. 2015; Lassaletta et al. 2016; Figure 1.1, Figure 1.3). Irrigation area
7 doubled, total nitrogen fertiliser use increased 9 times (FAOSTAT 2018; IFASTAT 2018) since the early
8 1960s. Biomass trade volumes grew by a factor of nine (in tons dry matter yr⁻¹) in this period, which is
9 much stronger than production (FAOSTAT 2018), resulting in a growing spatial disconnect between
10 regions of production and consumption (Friis et al. 2016; Friis and Nielsen 2017; Schröter et al. 2018; Liu
11 et al. 2013; Krausmann and Langthaler 2019). Urban and other infrastructure areas expanded by a factor 2
12 since 1960 (Krausmann et al. 2013), resulting in disproportionately large losses of highly-fertile cropland
13 (Seto and Reenberg 2014; Martellozzo et al. 2015; Bren d'Amour et al. 2016; Seto and Ramankutty 2016;
14 van Vliet et al. 2017). World regions show distinct patterns of change (Figure 1.3).

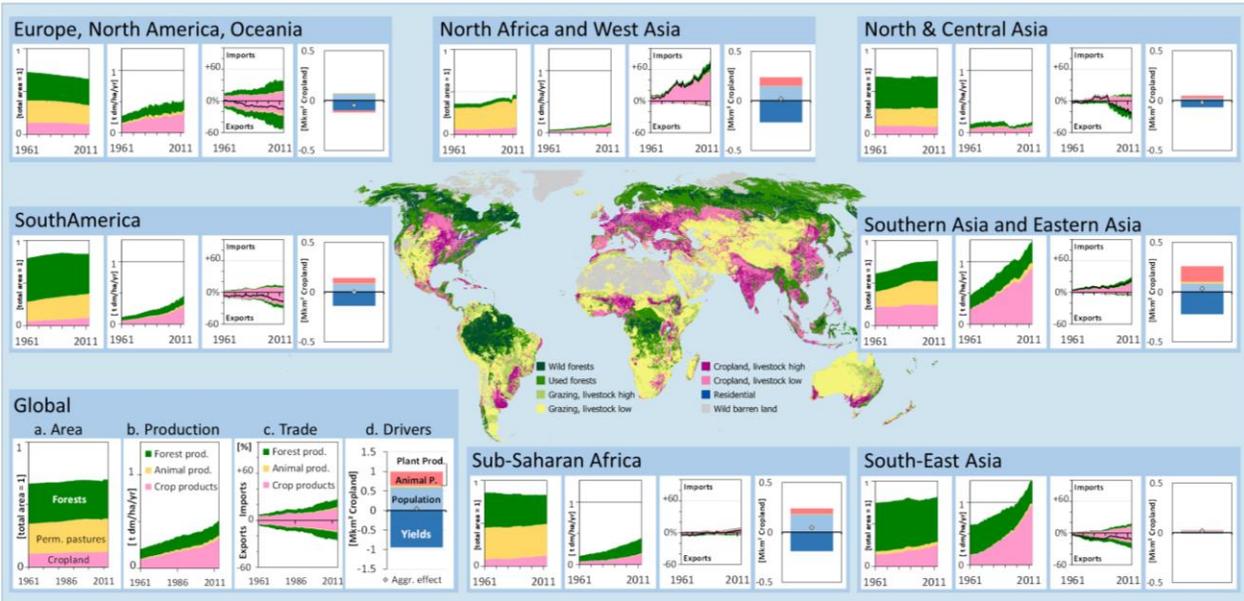
15 While most pastureland expansion replaced natural grasslands, cropland expansion replaced mainly
16 forests (Ramankutty et al. 2018; Ordway et al. 2017; Richards and Friess 2016). Noteworthy large
17 conversions occurred in tropical dry woodlands and savannahs, for example, in the Brazilian Cerrado
18 (Lehmann and Parr 2016; Strassburg et al. 2017), the South-American Caatinga and Chaco regions (Parr
19 et al. 2014; Lehmann and Parr 2016) or African savannahs (Ryan et al. 2016). More than half of the
20 original 4.3–12.6 million km² global wetlands (Erb et al. 2016a; Davidson 2014; Dixon et al. 2016) have
21 been drained; since 1970 the wetland extend index, developed by aggregating data field-site time series
22 that report changes in local wetland area indicate a decline by > 30% (Figure 1.1, see Section 4.2.1,
23 Darrah et al. 2019). Likewise, one third of the estimated global area that in a non-used state would be
24 covered in forests (Erb et al. 2017) has been converted to agriculture.

25 Global forest area declined by 3% since 1990 (about -5% since 1960) and continues to do so (FAO 2015a;
26 Keenan et al. 2015; MacDicken et al. 2015; FAO 1963; Figure 1.1), but uncertainties are large. *Low*
27 *agreement* relates to the concomitant trend of global tree-cover. Some remote-sensing based assessments
28 show global net-losses of forest or tree cover (Li et al. 2016; Nowosad et al. 2018; Hansen et al. 2013),
29 others indicate a net gain (Song et al. 2018). Tree-cover gains would be in line with observed and
30 modelled increases in photosynthetic active tissues (“greening”; Chen et al. 2019; Zhu et al. 2016; Zhao et
31 al. 2018; de Jong et al. 2013; Pugh et al. 2019; De Kauwe et al. 2016; Kolby Smith et al. 2015; see Box
32 2.3 in Chapter 2), but *confidence* remains *low* whether gross forest or tree cover gains are as large, or
33 larger, than losses. This uncertainty, together with poor information on forest management, affects
34 estimates and attribution of the land carbon sink (see Section 2.3, 4.3, 4.6). Discrepancies are caused by
35 different classification schemes and applied thresholds (e.g., minimum tree height and tree cover
36 thresholds used to define a forest), the divergence of forest and tree cover, and differences in methods and
37 spatiotemporal resolution (Keenan et al. 2015; Schepaschenko et al. 2015; Bastin et al. 2017; Sloan and
38 Sayer 2015; Chazdon et al. 2016a; Achard et al. 2014). However, there is *robust evidence and high*
39 *agreement* that a net loss of forest and tree cover prevails in the tropics and a net-gain, mainly of
40 secondary, semi-natural and planted, forests, in the temperate and boreal zones.

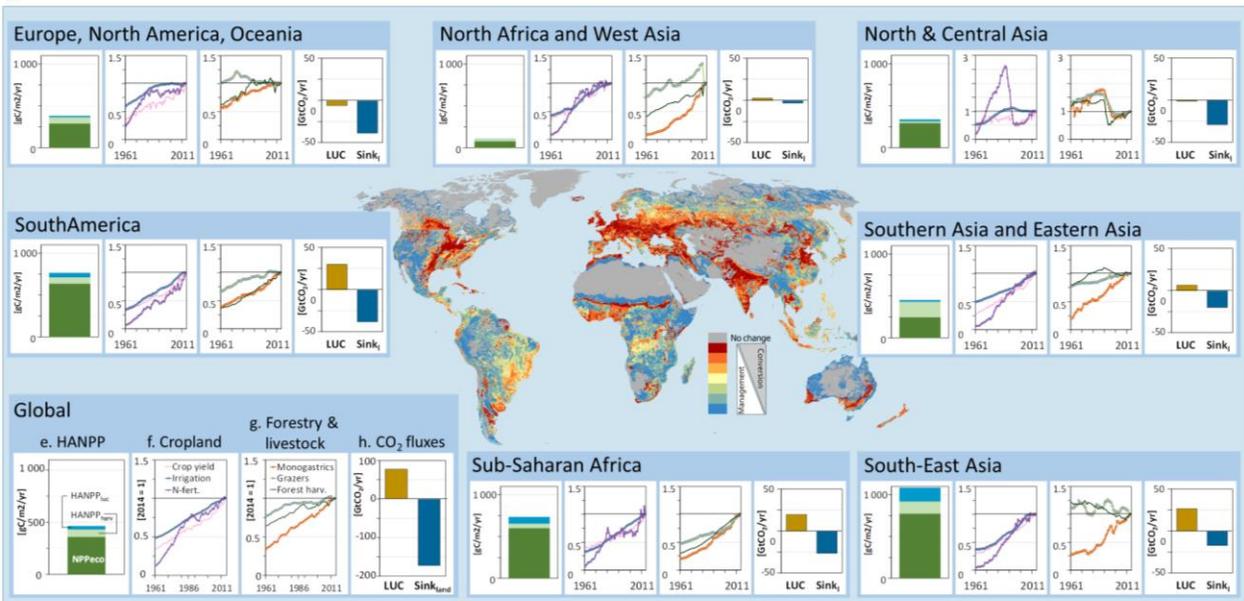
41 The observed regional and global historical land-use trends result in regionally distinct patterns of C
42 fluxes between land and the atmosphere (Figure 1.3B). They are also associated with declines in
43 biodiversity, far above background rates (Ceballos et al. 2015; De Vos et al. 2015; Pimm et al. 2014;
44 Newbold et al. 2015; Maxwell et al. 2016; Marques et al. 2019). Biodiversity losses from past global

1 land-use change have been estimated to be about 8–14%, depending on the biodiversity indicator applied
 2 (Newbold et al. 2015; Wilting et al. 2017; Gossner et al. 2016; Newbold et al. 2018; Paillet et al. 2010).
 3 In future, climate warming has been projected to accelerate losses of species diversity rapidly (Settele et
 4 al. 2014; Urban et al. 2016; Scholes et al. 2018; Fischer et al. 2018; Hoegh-Guldberg et al. 2018). The
 5 concomitance of land-use and climate-change pressures render ecosystem restoration a key challenge
 6 (Anderson-Teixeira 2018; Yang et al. 2019; see Section 4.8, 4.9).

A



B



7
 8 **Figure 1.3 Status and trends in the global land system. A. Trends in area, production and trade, and drivers**
 9 **of change. The map shows the global pattern of land systems (combination of maps Nachtergaele (2008); Ellis**
 10 **et al. (2010); Potapov et al. (2017); FAO’s Animal Production and Health Division (2018); livestock low/high**
 11 **relates to low or high livestock density, respectively). The inlay figures show, for the globe and 7 world**
 12 **regions, from left to right: (a) Cropland, permanent pastures and forest (used and unused) areas,**

1 standardised to total land area, (b) production in dry matter per year per total land area, (c) trade in dry
2 matter in percent of total domestic production, all for 1961 to 2014 (data from FAOSTAT (2018) and FAO
3 (1963) for forest area 1961). (d) drivers of cropland for food production between 1994 and 2011 (Alexander et
4 al. 2015). See panel “global” for legend. “Plant Produc., Animal P.”: changes in consumption of plant-based
5 products and animal-products, respectively. B. Selected land-use pressures and impacts. The map shows the
6 ratio between impacts on biomass stocks of land cover conversions and of land management (changes that
7 occur with land cover types; only changes larger than 30 gCm⁻² displayed; Erb et al. 2017), compared to the
8 biomass stocks of the potential vegetation (vegetation that would prevail in the absence of land use, but with
9 current climate). The inlay figures show, from left to right (e) the global Human Appropriation of Net
10 Primary production (HANPP) in the year 2005, in gCm⁻²yr⁻¹ (Krausmann et al. 2013). The sum of the three
11 components represents the NPP of the potential vegetation and consist of: (i) NPP_{eco}, i.e. the amount of NPP
12 remaining in ecosystem after harvest, (ii) HANPP_{harv}, i.e. NPP harvested or killed during harvest, and (iii)
13 HANPP_{luc}, i.e. NPP foregone due to land-use change. The sum of NPP_{eco} and HANPP_{harv} is the NPP of the
14 actual vegetation (Haberl et al. 2014; Krausmann et al. 2013). The two central inlay figures show changes in
15 land-use intensity, standardised to 2014, related to (f) cropland (yields, fertilisation, irrigated area) and (g)
16 forestry harvest per forest area, and grazers and monogastric livestock density per agricultural area
17 (FAOSTAT 2018). (h) Cumulative CO₂ fluxes between land and the atmosphere between 2000 and 2014.
18 LUC: annual CO₂ land use flux due to changes in land cover and forest management; Sink_{land}: the annual
19 CO₂ land sink caused mainly by the indirect anthropogenic effects of environmental change (e.g. climate
20 change and the fertilising effects of rising CO₂ and N concentrations), excluding impacts of land-use change
21 (Le Quéré et al. 2018; see Section 2.3).

22 1.2 Key challenges related to land use change

23 1.2.1 Land system change, land degradation, desertification and food security

24 1.2.1.1 Future trends in the global land system

25 Human population is projected to increase to nearly 9.8 (± 1) billion people by 2050 and 11.2 billion by
26 2100 (United Nations 2018). More people, a growing global middle class (Crist et al. 2017), economic
27 growth, and continued urbanisation (Jiang and O’Neill 2017) increase the pressures on expanding crop
28 and pasture area and intensifying land management. Changes in diets, efficiency and technology could
29 reduce these pressures (Billen et al. 2015; Popp et al. 2016; Muller et al. 2017; Alexander et al. 2015;
30 Springmann et al. 2018; Myers et al. 2017; Erb et al. 2016c; FAO 2018b; see Section 5.3, Section 6.2.2).

31 Given the large uncertainties underlying the many drivers of land use, as well as their complex relation to
32 climate change and other biophysical constraints, future trends in the global land system are explored in
33 scenarios and models that seek to span across these uncertainties (see Cross-Chapter Box 1: Scenarios, in
34 this Chapter). Generally, these scenarios indicate a continued increase in global food demand, owing to
35 population growth and increasing wealth. The associated land area needs are a key uncertainty, a function
36 of the interplay between production, consumption, yields, and production efficiency (in particular for
37 livestock and waste)(FAO 2018b; van Vuuren et al. 2017; Springmann et al. 2018; Riahi et al. 2017;
38 Prestele et al. 2016; Ramankutty et al. 2018; Erb et al. 2016b; Popp et al. 2016; see 1.3 and Cross-Chapter
39 Box 1: Scenarios, in this Chapter). Many factors, such as climate change, local contexts, education,
40 human and social capital, policy-making, economic framework conditions, energy availability,
41 degradation, and many more, affect this interplay, as discussed in all chapters of this report.

42 Global telecouplings in the land system, the distal connections and multidirectional flows between
43 regions and land systems, are expected to increase, due to urbanisation (Seto et al. 2012; van Vliet et al.
44 2017; Jiang and O’Neill 2017; Friis et al. 2016), and international trade (Konar et al. 2016; Erb et al.
45 2016b; Billen et al. 2015; Lassaletta et al. 2016). Telecoupling can support efficiency gains in production,
46 but can also lead to complex cause-effect chains and indirect effects such as land competition or leakage
47 (displacement of the environmental impacts, see glossary), with governance challenges (Baldos and

1 Hertel 2015; Kastner et al. 2014; Liu et al. 2013; Wood et al. 2018; Schröter et al. 2018; Lapola et al.
2 2010; Jadin et al. 2016; Erb et al. 2016b; Billen et al. 2015; Chaudhary and Kastner 2016; Marques et al.
3 2019; Seto and Ramankutty 2016; see Section 1.2.1.5). Furthermore, urban growth is anticipated to occur
4 at the expense of fertile (crop)land, posing a food security challenge, in particular in regions of high
5 population density and agrarian-dominated economies, with limited capacity to compensate for these
6 losses (Seto et al. 2012; Güneralp et al. 2013; Aronson et al. 2014; Martellozzo et al. 2015; Bren d'Amour
7 et al. 2016; Seto and Ramankutty 2016; van Vliet et al. 2017).

8 Future climate change and increasing atmospheric CO₂ concentration are expected to accentuate existing
9 challenges by, for example, shifting biomes or affecting crop yields (Baldos and Hertel 2015; Schlenker
10 and Lobell 2010; Lipper et al. 2014; Challinor et al. 2014; Myers et al. 2017; see Section 5.2.2), as well
11 as through land-based, climate change mitigation. There is *high confidence* that large-scale
12 implementation of bioenergy or afforestation can further exacerbate existing challenges (Smith et al. 2016;
13 see also Section 1.3.1 and Cross-chapter box 7 on bioenergy in Chapter 6).

14 **1.2.1.2 Land Degradation**

15 As discussed in Chapter 4, the concept of land degradation, including its definition, has been used in
16 different ways in different communities and in previous assessments (such as the IPBES Land
17 degradation and restoration assessment). In the SRCCL, land degradation is defined as a *negative trend in*
18 *land condition, caused by direct or indirect human-induced processes including anthropogenic climate*
19 *change, expressed as long-term reduction or loss of at least one of the following: biological productivity,*
20 *ecological integrity or value to humans.* This definition applies to forest and non-forest land (see Chapter
21 4 and Glossary).

22 Land degradation is a critical issue for ecosystems around the world due to the loss of actual or potential
23 productivity or utility (Ravi et al. 2010; Mirzabaev et al. 2015; FAO and ITPS 2015; Cerretelli et al.
24 2018). Land degradation is driven to a large degree by unsustainable agriculture and forestry,
25 socioeconomic pressures, such as rapid urbanisation and population growth, and unsustainable production
26 practices in combination with climatic factors (Field et al. 2014b; Lal 2009; Beinroth, F. H., Eswaran, H.,
27 Reich, P. F. and Van Den Berg 1994; Abu Hammad and Tumeizi 2012; Ferreira et al. 2018; Franco and
28 Giannini 2005; Abahussain et al. 2002).

29 Global estimates of the total degraded area (excluding deserted area) vary from less than 10 million km²
30 to over 60 million km², with additionally large disagreement regarding the spatial distribution (Gibbs and
31 Salmon 2015; see Section 4.3). The annual increase in the degraded land area has been estimated as
32 50,000–10,000 million km² yr⁻¹ (Stavi and Lal 2015), and the loss of total ecosystem services equivalent
33 to about 10% of the world's GDP in the year 2010 (Sutton et al. 2016). Although land degradation is a
34 common risk across the globe, poor countries remain most vulnerable to its impacts. Soil degradation is
35 of particular concern, due to the long period necessary to restore soils (Lal 2009; Stockmann et al. 2013;
36 Lal 2015), as well as the rapid degradation of primary forests through fragmentation (Haddad et al. 2015).
37 Among the most vulnerable ecosystems to degradation are high carbon stock wetlands (including
38 peatlands). Drainage of natural wetlands for use in agriculture leads to high CO₂ emissions and
39 degradation (*high confidence*) (Strack 2008; Limpens et al. 2008; Aich et al. 2014; Murdiyarso et al.
40 2015; Kauffman et al. 2016; Dohong et al. 2017; Arifanti et al. 2018; Evans et al. 2019). Land
41 degradation is an important factor contributing to uncertainties in the mitigation potential of land-based
42 ecosystems (Smith et al. 2014). Furthermore, degradation that reduces forest (and agricultural) biomass
43 and soil organic carbon leads to higher rates of runoff (*high confidence*) (Molina et al. 2007; Valentin et

1 al. 2008; Mateos et al. 2017; Noordwijk et al. 2017) and hence to increasing flood risk (*low confidence*)
2 (Bradshaw et al. 2007; Laurance 2007; van Dijk et al. 2009).

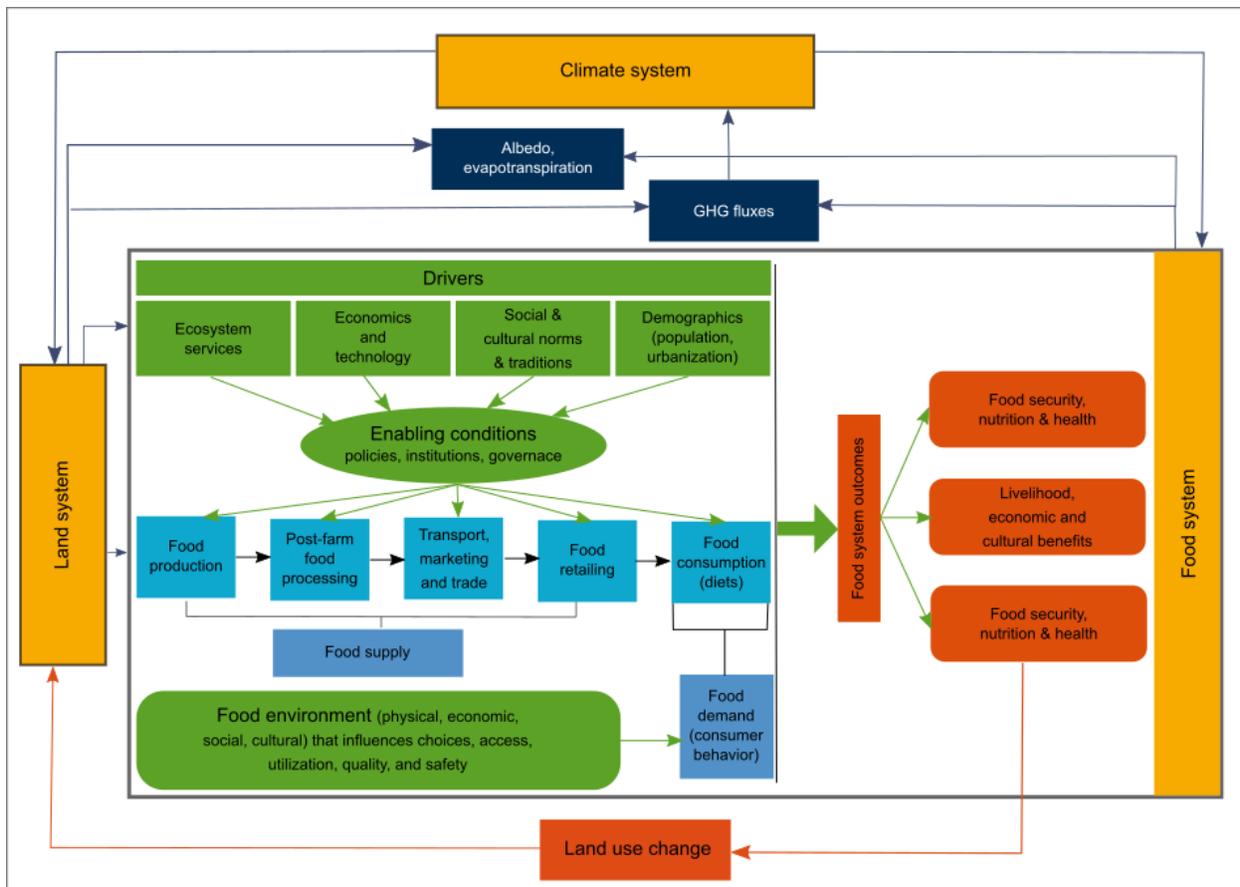
3 **1.2.1.3 Desertification**

4 The SRCCL adopts the definition of the UNCCD of desertification being land degradation in arid, semi-
5 arid and dry sub-humid areas (drylands) (see glossary, and Section 3.1.1). Desertification results from
6 various factors, including climate variations and human activities, and is not limited to irreversible forms
7 of land degradation (Tal 2010)(Bai et al. 2008). A critical challenge in the assessment of desertification is
8 to identify a “non-desertified” reference state (Bestelmeyer et al. 2015). While climatic trends and
9 variability can change the intensity of desertification processes, some authors exclude climate effects,
10 arguing that desertification is a purely human-induced process of land degradation with different levels of
11 severity and consequences (Sivakumar 2007).

12 As a consequence of varying definitions and different methodologies, the area of desertification varies
13 widely (see (D’Odorico et al. 2013; Bestelmeyer et al. 2015), and references therein). Arid regions of the
14 world cover up to about 46% of the total terrestrial surface (about 60 million km²; Pravalie 2016;
15 Koutroulis 2019). Around 3 billion people reside in dryland regions (D’Odorico et al. 2013; Maestre et al.
16 2016; see Section 3.1.1), and the number of people living in areas affected by desertification has been
17 estimated as > 630 million, compared to 211 million in the early 1960s (see Fig. 1.1, see Section 3.1.1).
18 The combination of low rainfall with frequently infertile soils renders these regions, and the people who
19 rely on them, vulnerable to both climate change, and unsustainable land management (*high confidence*).
20 In spite of the national, regional and international efforts to combat desertification, it remains one of the
21 major environmental problems (Abahussain et al. 2002; Cherlet et al. 2018).

22 **1.2.1.4 Food security, food systems and linkages to land-based ecosystems**

23 The High Level Panel of Experts of the Committee on Food Security define the food system as to “gather
24 all the elements (environment, people, inputs, processes, infrastructures, institutions, etc.) and activities
25 that relate to the production, processing, distribution, preparation and consumption of food, and the
26 output of these activities, including socio-economic and environmental outcomes” (HLPE 2017).
27 Likewise, food security has been defined as “a situation that exists when all people, at all times, have
28 physical, social and economic access to sufficient, safe and nutritious food that meets their dietary needs
29 and food preferences for an active and healthy life “ (FAO 2017). By this definition, food security is
30 characterised by food availability, economic and physical access to food, food utilisation and food
31 stability over time. Food and nutrition security is one of the key outcomes of the food system (FAO
32 2018b; Figure 1.4).



1
 2 **Figure 1.4 Food system (and its relations to land and climate):** The food system is conceptualised through
 3 **supply (production, processing, marketing and retailing) and demand (consumption and diets) that are**
 4 **shaped by physical, economic, social and cultural determinants influencing choices, access, utilisation,**
 5 **quality, safety and waste. Food system drivers (ecosystem services, economics and technology, social and**
 6 **cultural norms and traditions, and demographics) combine with the enabling conditions (policies, institutions**
 7 **and governance) to affect food system outcomes including food security, nutrition and health, livelihoods,**
 8 **economic and cultural benefits as well as environmental outcomes or side-effects (nutrient and soil loss, water**
 9 **use and quality, GHG emissions and other pollutants). Climate and climate change has direct impact on the**
 10 **food system (productivity, variability, nutritional quality) while the latter contribute to local climate (albedo,**
 11 **evapotranspiration) and global warming (GHGs). The land system (function, structures, and processes) affect**
 12 **the food system directly (food production) and indirectly (ecosystem services) while food demand and supply**
 13 **processes affect land (land use change) and land-related processes (e.g., land degradation, desertification) (see**
 14 **chapter 5).**

15 After a prolonged decline, world hunger appears to be on the rise again with the number of
 16 undernourished people having increased to an estimated 821 million in 2017, up from 804 million in 2016
 17 and 784 million in 2015, although still below the 900 million reported in 2000 (FAO et al. 2018; see
 18 Section 5.1.2). Of the total undernourished in 2018, lived, for example, 256.5 million in Africa, and 515.1
 19 million in Asia (excluding Japan). The same report also states that child undernourishment continues to
 20 decline, but levels of overweight populations and obesity are increasing. The total number of overweight
 21 children in 2017 was 38-40 million worldwide, and globally up to around two billion adults are by now
 22 overweight (see Section 5.1.2). FAO also estimated that close to 2000 million people suffer from
 23 micronutrient malnutrition (FAO 2018b).

1 Food insecurity most notably occurs in situations of conflict and conflict combined with droughts or
2 floods (Cafiero et al. 2018; Smith et al. 2017). The close parallel between food insecurity prevalence and
3 poverty means that tackling development priorities would enhance sustainable land use options for
4 climate mitigation.

5 Climate change affects the food system as changes in trends and variability in rainfall and temperature
6 variability impact crop and livestock productivity and total production (Osborne and Wheeler 2013;
7 Tigchelaar et al. 2018; Iizumi and Ramankutty 2015), the nutritional quality of food (Loladze 2014;
8 Myers et al., 2014; Ziska et al. 2016; Medek et al., 2017), water supply (Nkhonjera 2017), and incidence
9 of pests and diseases (Curtis et al. 2018). These factors also impact on human health and increase
10 morbidity and affect human ability to process ingested food (Franchini and Mannucci 2015; Wu et al.
11 2016; Raiten and Aimone 2017). At the same time, the food system generates negative externalities (the
12 environmental effects of production and consumption) in the form of GHG emissions (Section 1.1.2,
13 Section 2.3), pollution (van Noordwijk and Brussaard 2014; Thyberg and Tonjes 2016; Borsato et al.
14 2018; Kibler et al. 2018), water quality (Malone et al. 2014; Norse and Ju 2015), and ecosystem services
15 loss (Schipper et al. 2014; Eraerts et al. 2017) with direct and indirect impacts on climate change and
16 reduced resilience to climate variability. As food systems are assessed in relation to their contribution to
17 global warming and/or to land degradation (e.g., livestock systems) it is critical to evaluate their
18 contribution to food security and livelihoods and to consider alternatives, especially for developing
19 countries where food insecurity is prevalent (Röös et al. 2017; Salmon et al. 2018).

20 **1.2.1.5 Challenges arising from land governance**

21 Land use change has both positive and negative effects: it can lead to economic growth, but it can become
22 a source of tension and social unrest leading to elite capture, and competition (Haberl 2015). Competition
23 for land plays out continuously among different use types (cropland, pastureland, forests, urban spaces,
24 and conservation and protected lands) and between different users within the same land use category
25 (subsistence vs. commercial farmers)(Dell'Angelo et al. 2017b). Competition is mediated through
26 economic and market forces (expressed through land rental and purchases, as well as trade and
27 investments). In the context of such transactions, power relations often disfavour disadvantaged groups
28 such as small scale farmers, indigenous communities or women (Doss et al. 2015; Ravnborg et al. 2016).
29 These drivers are influenced to a large degree by policies, institutions and governance structures. Land
30 governance determines not only who can access the land, but also the role of land ownership (legal,
31 formal, customary or collective) which influences land use, land use change and the resulting land
32 competition (Moroni 2018).

33 Globally, there is competition for land because it is a finite resource and because most of the highly-
34 productive land is already exploited by humans (Lambin and Meyfroidt 2011; Lambin 2012; Venter et al.
35 2016). Driven by growing population, urbanisation, demand for food and energy, as well as land
36 degradation, competition for land is expected to accentuate land scarcity in the future(Tilman et al. 2011;
37 Foley et al. 2011; Lambin 2012; Popp et al. 2016)(*robust evidence, high agreement*). Climate change
38 influences land use both directly and indirectly, as climate policies can also play a role in increasing
39 land competition via forest conservation policies, afforestation, or energy crop production (see Section
40 1.3.1), with the potential for implications for food security (Hussein et al. 2013) and local land-ownership.

41 An example of large-scale change in land ownership is the much-debated large-scale land acquisition
42 (LSLA) by investors which peaked in 2008 during the food price crisis, the financial crisis, and has also
43 been linked to the search for biofuel investments (Dell'Angelo et al. 2017a). Since 2000, almost 50
44 million hectares of land, have been acquired, and there are no signs of stagnation in the foreseeable future

1 (Land Matrix 2018). The LSLA phenomenon, which largely targets agriculture, is widespread, including
 2 Sub-Saharan Africa, Southeast Asia, Eastern Europe and Latin America (Rulli et al. 2012; Nolte et al.
 3 2016; Constantin et al. 2017). LSLAs are promoted by investors and host governments on economic
 4 grounds (infrastructure, employment, market development)(Deininger et al. 2011), but their social and
 5 environmental impacts can be negative and significant (Dell’Angelo et al. 2017a).

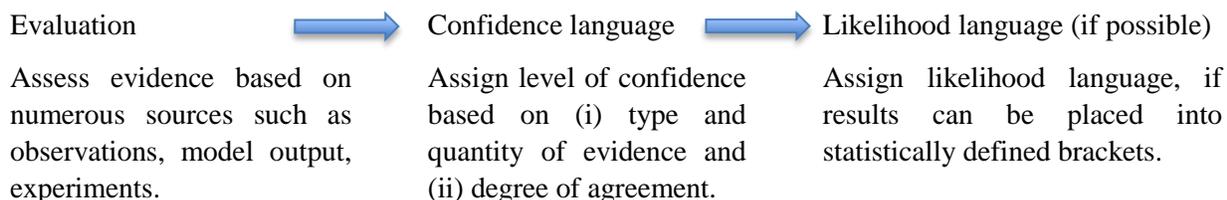
6 Much of the criticism of LSLA focuses on their social impacts, especially the threat to local communities’
 7 land rights (especially indigenous people and women) (Anseeuw et al. 2011) and displaced communities
 8 creating secondary land expansion (Messerli et al. 2014; Davis et al. 2015). The promises that LSLAs
 9 would develop efficient agriculture on non-forested, unused land (Deininger et al. 2011) has so far not
 10 been fulfilled. However, LSLAs is not the only outcome of weak land governance structures (Wang et al.
 11 2016), other forms of inequitable or irregular land acquisition can also be home-grown pitting one
 12 community against a more vulnerable group (Xu 2018) or land capture by urban elites (McDonnell 2017).
 13 As demands on land are increasing, building governance capacity and securing land tenure becomes
 14 essential to attain sustainable land use, which has the potential to mitigate climate change, promote food
 15 security, and potentially reduce risks of climate-induced migration and associated risks of conflicts (see
 16 Section 7.6).

17 **1.2.2 Progress in dealing with uncertainties in assessing land processes in the climate** 18 **system**

19 **1.2.2.1 Concepts related to risk, uncertainty and confidence**

20 In context of the SRCCL, risk refers to the potential for the adverse consequences for human or (land-
 21 based) ecological systems, arising from climate change or responses to climate change. Risk related to
 22 climate change impacts integrates across the hazard itself, the time of exposure and the vulnerability of
 23 the system; the assessment of all three of these components, their interactions, and outcomes are uncertain
 24 (see glossary for expanded definition and Section 7.1.2). For instance, a risk to human society is the
 25 continued loss of productive land which might arise from climate change, mismanagement, or a
 26 combination of both factors. However, risk can also arise from the potential for adverse consequences
 27 from responses to climate change, such as widespread deployment of bioenergy which is intended to
 28 reduce greenhouse gas emissions and thus limit climate change, but can present its own risks to food
 29 security (see chapters 5, 6 and 7).

30 Demonstrating with some statistical certainty that the climate or the land system affected by climate or
 31 land use has changed (detection), and evaluating the relative contributions of multiple causal factors to
 32 that change (with a formal assessment of confidence; attribution. See glossary) remain challenging
 33 aspects in both observations and models (Rosenzweig and Neofotis 2013; Gillett et al. 2016; Lean 2018).
 34 Uncertainties arising for example, from missing or imprecise data, ambiguous terminology, incomplete
 35 process representation in models, or human decision making contribute to these challenges, and some
 36 examples are provided in this subsection. In order to reflect various sources of uncertainties in the state of
 37 scientific understanding, IPCC assessment reports provide estimates of confidence (Mastrandrea et al.
 38 2011). This confidence language is also used in the SRCCL (Figure 1.5):



| | | | | |
|-------------|---|-------------------------------------|-------------------------------------|------------------------------|
| Agreement ↑ | High agreement Limited evidence | High agreement Medium evidence | High agreement Robust evidence | Confidence low high |
| | Medium agreement Limited evidence | Medium agreement Medium evidence | Medium agreement Robust evidence | |
| | Low agreement Limited evidence | Low agreement Medium evidence | Low agreement Robust evidence | |
| | Evidence (type, amount, quality, consistency) → | | | |

Figure 1.5 Use of confidence language

1.2.2.2 Nature and scope of uncertainties related to land use

Identification and communication of uncertainties is crucial to support decision making towards sustainable land management. Providing a robust, and comprehensive understanding of uncertainties in observations, models and scenarios is a fundamental first step in the IPCC confidence framework (see above). This will remain a challenge in future, but some important progress has been made over recent years.

Uncertainties in observations

The detection of changes in vegetation cover and structural properties underpins the assessment of land-use change, degradation and desertification. It is continuously improving by enhanced Earth observation capacity (Hansen et al. 2013; He et al. 2018; Ardö et al. 2018; Spennemann et al. 2018) (see also Table SM. 1.1 in Supplementary Materials). Likewise, the picture of how soil organic carbon, and GHG and water fluxes respond to land-use change and land management continues to improve through advances in methodologies and sensors (Kostyanovsky et al. 2018; Brümmer et al. 2017; Iwata et al. 2017; Valayamkunnath et al. 2018). In both cases, the relative shortness of the record, data gaps, data treatment algorithms and –for remote sensing- differences in the definitions of major vegetation cover classes limits the detection of trends (Alexander et al. 2016a; Chen et al. 2014; Yu et al. 2014; Lacaze et al. 2015; Song 2018; Peterson et al. 2017). In many developing countries, the cost of satellite remote sensing remains a challenge, although technological advances are starting to overcome this problem (Santilli et al. 2018), while ground-based observations networks are often not available.

Integration of multiple data sources in model and data assimilation schemes reduces uncertainties (Li et al. 2017; Clark et al. 2017; Lees et al. 2018), which might be important for the advancement of early warning systems. Early warning systems are a key feature of short-term (i.e. seasonal) decision support systems and are becoming increasingly important for sustainable land management and food security (Shtienberg 2013; Jarroudi et al. 2015; see Section 6.2.3, 7.4.3). Early warning systems can help to optimise fertiliser and water use, aid disease suppression, and/or increase the economic benefit by enabling strategic farming decisions on when and what to plant (Caffi et al. 2012; Watmuff et al. 2013; Jarroudi et al. 2015; Chipanshi et al. 2015). Their suitability depends on the capability of the methods to accurately predict crop or pest developments, which in turn depends on expert agricultural knowledge, and the accuracy of the weather data used to run phenological models (Caffi et al. 2012; Shtienberg 2013).

Uncertainties in models

Model intercomparison is a widely used approach to quantify some sources of uncertainty in climate change, land-use change and ecosystem modelling, often associated with the calculation of model-ensemble medians or means (see e.g., Section 2.2; Section 5.2). Even models of broadly similar structure differ in their projected outcome for the same input, as seen for instance in the spread in climate change

1 projections from Earth System Models (ESMs) to similar future anthropogenic GHG emissions (Parker
2 2013; Stocker et al. 2013a). These uncertainties arise, for instance, from different parameter values,
3 different processes represented in models, or how these processes are mathematically described. If the
4 output of ESM simulations are used as input to impact models, these uncertainties can propagate to
5 projected impacts (Ahlstrom et al. 2013).

6 Thus, the increased quantification of model performance in benchmarking exercises (the repeated
7 confrontation of models with observations to establish a track-record of model developments and
8 performance) is an important development to support the design and the interpretation of the outcomes of
9 model ensemble studies (Randerson et al. 2009; Luo et al. 2012; Kelley et al. 2013). Since observational
10 data sets in themselves are uncertain, benchmarking benefits from transparent information on the
11 observations that are used, and the inclusion of multiple, regularly updated data sources (Luo et al. 2012;
12 Kelley et al. 2013). Improved benchmarking approaches and the associated scoring of models may
13 support weighted model means contingent on model performance. This could be an important step
14 forward when calculating ensemble means across a range of models (Buisson et al. 2009; Parker 2013;
15 Prestele et al. 2016).

16 *Uncertainties arising from unknown futures*

17 Large differences exist in projections of future land cover change, both between and within scenario
18 projections (Fuchs et al. 2015; Eitelberg et al. 2016; Popp et al. 2016; Krause et al. 2017; Alexander et al.
19 2016a). These differences reflect the uncertainties associated with baseline data, thematic classifications,
20 different model structures and model parameter estimation (Alexander et al. 2017a; Prestele et al. 2016;
21 Cross-Chapter Box 1: Scenarios, in this Chapter). Likewise, projections of future land-use change are also
22 highly uncertain, reflecting –among other factors- the absence of important crop, pasture and management
23 processes in Integrated Assessment Models (Cross-Chapter Box 1: Scenarios, in this Chapter; Rose 2014)
24 and in models of the terrestrial carbon cycle (Arneth et al. 2017). These processes have been shown to
25 have large impacts on carbon stock changes (Arneth et al. 2017). Common scenario frameworks are used
26 to capture the range of future uncertainties in scenarios. The most commonly used recent framework in
27 climate change studies is based on the Representative Concentration Pathways (RCPs) and the Shared
28 Socio-economic Pathways (SSPs)(Popp et al. 2016; Riahi et al. 2017). The RCPs prescribe levels of
29 radiative forcing (Wm^{-2}) arising from different atmospheric concentrations of GHGs that lead to different
30 levels of climate change. For example, RCP2.6 (2.6 Wm^{-2}) is projected to lead to global mean
31 temperature changes of about 0.9°C – 2.3°C , and RCP8.5 (8.5 Wm^{-2}) to global mean temperature changes
32 of about 3.2°C – 5.4°C (van Vuuren et al 2014).

33 The SSPs describe alternative trajectories of future socio-economic development with a focus on
34 challenges to climate mitigation and challenges to climate adaptation (O’Neill et al. 2014). SSP1
35 represents a sustainable and co-operative society with a low carbon economy and high capacity to adapt
36 to climate change. SSP3 has social inequality that entrenches reliance on fossil fuels and limits adaptive
37 capacity. SSP4 has large differences in income within and across world regions that facilitates low carbon
38 economies in places, but limits adaptive capacity everywhere. SSP5 is a technologically advanced world
39 with a strong economy that is heavily dependent on fossil fuels, but with high adaptive capacity. SSP2 is
40 an intermediate case between SSP1 and SSP3 (O’Neill et al. 2014). The SSPs are commonly used with
41 models to project future land use change (Cross-Chapter Box 1: Scenarios, in this Chapter).

42 The SSPs map onto the RCPs through shared assumptions. For example, a higher level of climate change
43 (RCP8.5) is associated with higher challenges for climate change mitigation (SSP5). Not all SSPs are,
44 however, associated with all RCPs. For example, an SSP5 world is committed to high fossil fuel use,

1 associated GHG emissions, and this is not commensurate with lower levels of climate change (e.g.,
2 RCP2.6). (Engstrom et al. 2016) took this approach further by ascribing levels of probability that
3 associate an SSP with an RCP, contingent on the SSP scenario assumptions (see Cross-Chapter Box 1:
4 Scenarios, in this Chapter).

Cross-Chapter Box 1: Scenarios and other methods to characterise the future of land

Mark Rounsevell (United Kingdom/Germany), Almut Arneth (Germany), Katherine Calvin (The United States of America), Edouard Davin (France/Switzerland), Jan Fuglestad (Norway), Joanna House (United Kingdom), Alexander Popp (Germany), Joana Portugal Pereira (United Kingdom), Prajal Pradhan (Nepal/Germany), Jim Skea (United Kingdom), David Viner (United Kingdom).

About this box

The land-climate system is complex and future changes are uncertain, but methods exist (collectively known as *futures analysis*) to help decision makers in navigating through this uncertainty. Futures analysis comprises a number of different and widely used methods, such as scenario analysis (Rounsevell and Metzger 2010), envisioning or target setting (Kok et al. 2018), pathways analysis¹ (IPBES 2016; IPCC 2018), and conditional probabilistic futures (Vuuren et al. 2018; Engstrom et al. 2016; Henry et al. 2018)(see Cross-Chapter Box 1, Table 1). Scenarios and other methods to characterise the future can support a discourse with decision makers about the sustainable development options that are available to them. All chapters of this assessment draw conclusions from futures analysis and so, the purpose of this box is to outline the principal methods used, their application domains, their uncertainties and their limitations.

Exploratory scenario analysis

Many exploratory scenarios are reported in climate and land system studies on climate change (Dokken 2014), land-based, climate-change mitigation for example, reforestation/afforestation, avoided deforestation and bioenergy (Kraxner et al. 2013; Humpenoder et al. 2014; Krause et al. 2017) and climate change impacts and adaptation (Warszawski et al. 2014). There are global-scale scenarios of food security (Foley et al. 2011; Pradhan et al. 2013, 2014), but fewer scenarios of desertification, land degradation and restoration (Wolff et al. 2018). Exploratory scenarios combine qualitative ‘storylines’ or descriptive narratives of the underlying causes (or drivers) of change (Nakicenovic and Swart 2000; Rounsevell and Metzger 2010; O’Neill et al. 2014) with quantitative projections from computer models. Different types of models are used for this purpose based on very different modelling paradigms, baseline data and underlying assumptions (Alexander et al. 2016a; Prestele et al. 2016). Cross-Chapter Box 1, Figure 1 outlines how a combination of models can quantify these components as well as the interactions between them.

Exploratory scenarios often show that socio-economic drivers have a larger effect on land use change than climate drivers (Harrison et al. 2014, 2016). Of these, technological development is critical in affecting the production potential (yields) of food and bioenergy and the feed conversion efficiency of livestock (Rounsevell et al. 2006; Wise et al. 2014; Kreidenweis et al. 2018), as well as the area of land needed for food production (Foley et al. 2011; Weindl et al. 2017; Kreidenweis et al. 2018). Trends in consumption, for example, diets, waste reduction, are also fundamental in affecting land use change (Pradhan et al. 2013; Alexander et al. 2016b; Weindl et al. 2017; Alexander et al. 2017; Vuuren et al.

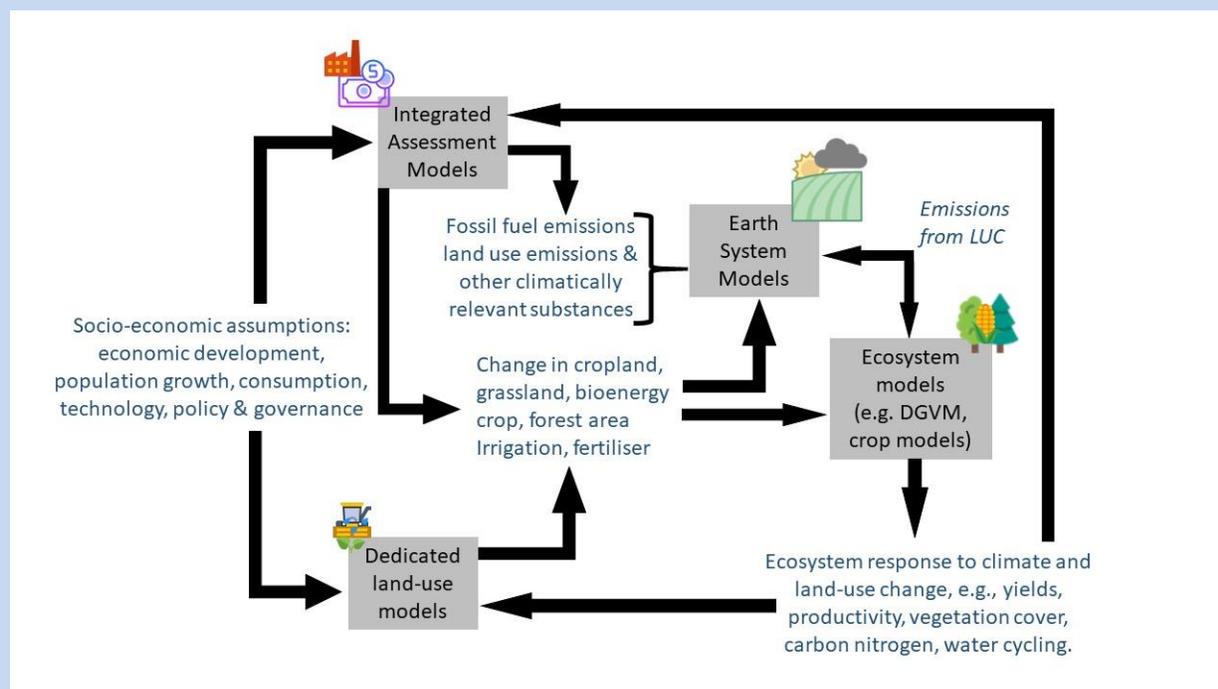
¹ FOOTNOTE: Different communities have a different understanding of the concept of pathways, as noted in the Cross-Chapter Box 1 on scenarios in (IPCC 2018). Here, we refer to pathways as a description of the time-dependent actions required to move from today’s world to a set of future visions (IPCC 2018). However, the term pathways is commonly used in the climate change literature as a synonym for projections or trajectories (e.g. Shared socio-economic pathways).

2018; Bajželj et al. 2014). Scenarios of land-based mitigation through large-scale bioenergy production and afforestation often lead to negative trade-offs with food security (food prices), water resources and biodiversity (cross chapter box on bioenergy, Ch6).

Cross-Chapter Box 1, Table 1 Description of the principal methods used in land and climate futures analysis

| Futures method | Description and subtypes | Application domain | Time horizon | Examples in this assessment |
|--|--|--|---|---|
| <i>Exploratory scenarios.</i> Trajectories of change in system components from the present to contrasting, alternative futures based on plausible and internally consistent assumptions about the underlying drivers of change | <i>Long-term projections</i> quantified with models | Climate system, land system and other components of the environment (e.g., biodiversity, ecosystem functioning, water resources and quality), for example the SSPs | 10-100 years | 2.3, 2.6.2, 5.2.3, 6.1.4, 6.4.4, 7.2 |
| | <i>Business-as-usual scenarios</i> (including 'outlooks') | A continuation into the future of current trends in key drivers to explore the consequences of these in the near-term | 5-10 years, 20-30 years for outlooks | 1.2.1, 2.6.2, 5.3.4, 6.1.4 |
| | <i>Policy & planning scenarios</i> (including business planning) | Ex Ante analysis of the consequences of alternative policies or decisions based on known policy options or already implemented policy and planning measures | 5-30 years | 2.6.3, 5.5.2, 5.6.2, 6.4.4 |
| | <i>Stylised scenarios</i> (with single and multiple options) | Afforestation/reforestation areas, bioenergy areas, protected areas for conservation, consumption patterns (e.g., diets, food waste) | 10-100 years | 2.6.1, 5.5.1, 5.5.2, 5.6.1, 5.6.2, 6.4.4, 7.2 |
| | <i>Shock scenarios</i> (high impact single events) | Food supply chain collapses, cyberattacks, pandemic diseases (humans, crops and livestock) | Near-term events (up to 10 years) leading to long-term impacts (10-100 years) | 5.8.1 |
| | <i>Conditional probabilistic futures</i> ascribe probabilities to uncertain drivers that are conditional on scenario assumptions | Where some knowledge is known about driver uncertainties, for example, population, economic growth, land use change | 10-100 years | 1.2 |
| <i>Normative scenarios.</i> Desired futures or outcomes that are aspirational and how to achieve them | <i>Visions, goal-seeking or target-seeking scenarios</i> | Environmental quality, societal development, human well-being, the Representative Concentration Pathways (RCPs,) 1.5 °C scenarios | 5-10 years to 10-100 years | 2.6.2, 6.4.4, 7.2, 5.5.2 |

| | | | | |
|--|---|---|----------------------------|-------------------|
| | <i>Pathways</i> as alternative sets of choices, actions or behaviours that lead to a future vision (goal or target) | Socio-economic systems, governance and policy actions | 5-10 years to 10-100 years | 5.5.2, 6.4.4, 7.2 |
|--|---|---|----------------------------|-------------------|



Cross-Chapter Box 1, Figure 1 Interactions between land and climate system components and models in scenario analysis. The blue text describes selected model inputs and outputs.

Many exploratory scenarios are based on common frameworks such as the Shared Socio-economic Pathways (SSPs) (Popp et al. 2016; Riahi et al. 2017; Doelman et al. 2018) (see section 1.2). However, other methods are used. *Stylised scenarios* prescribe assumptions about climate and land use change solutions for example, dietary change, food waste reduction, afforestation areas (Pradhan et al. 2013, 2014; Kreidenweis et al. 2016; Rogelj et al. 2018b; Seneviratne et al. 2018; Vuuren et al. 2018). These scenarios provide useful thought experiments, but the feasibility of achieving the stylised assumptions is often unknown. *Shock scenarios* explore the consequences of low probability, high impact events such as pandemic diseases, cyberattacks and failures in food supply chains (Challinor et al. 2018) often in food security studies. Because of the diversity of exploratory scenarios, attempts have been made to categorise them into ‘archetypes’ based on the similarity between their assumptions in order to facilitate communication (IPBES 2018a).

Conditional probabilistic futures explore the consequences of model parameter uncertainty in which these uncertainties are conditional on scenario assumptions (Neill 2004). Only a few studies have applied the conditional probabilistic approach to land use futures (Brown et al. 2014; Engstrom et al. 2016; Henry et al. 2018). By accounting for uncertainties in key drivers these studies show large ranges in land use change, for example, global cropland areas of 893–2380 Mha by the end of the 21st Century (Engstrom et al. 2016). They also find that land-use targets may not be achieved, even across a wide

range of scenario parameter settings, because of trade-offs arising from the competition for land (Henry et al. 2018; Heck et al. 2018). Accounting for uncertainties across scenario assumptions can lead to convergent outcomes for land use change, which implies that certain outcomes are more robust across a wide range of uncertain scenario assumptions (Brown et al. 2014).

In addition to global scale scenario studies, sub-national studies demonstrate that regional climate change impacts on the land system are highly variable geographically because of differences in the spatial patterns of both climate and socio-economic change (Harrison et al. 2014). Moreover, the capacity to adapt to these impacts is strongly dependent on the regional, socio-economic context and coping capacity (Dunford et al. 2014); processes that are difficult to capture in global scale scenarios. Regional scenarios are often co-created with stakeholders through participatory approaches (Kok et al. 2014), which is powerful in reflecting diverse worldviews and stakeholder values. Stakeholder participatory methods provide additional richness and context to storylines, as well as providing saliency and legitimacy for local stakeholders (Kok et al. 2014).

Normative scenarios: visions and pathways analysis

Normative scenarios reflect a desired or target-seeking future. Pathways analysis is important in moving beyond the ‘*what if?*’ perspective of exploratory scenarios to evaluate how normative futures might be achieved in practice, recognising that multiple pathways may achieve the same future vision. Pathways analysis focuses on consumption and behavioural changes through transitions and transformative solutions (IPBES 2018a). Pathways analysis is highly relevant in support of policy, since it outlines sets of time-dependent actions and decisions to achieve future targets, especially with respect to sustainable development goals, as well as highlighting trade-offs and co-benefits (IPBES 2018a). Multiple, alternative pathways have been shown to exist that mitigate trade-offs whilst achieving the priorities for future sustainable development outlined by governments and societal actors. Of these alternatives, the most promising focus on long-term societal transformations through education, awareness raising, knowledge sharing and participatory decision-making (IPBES 2018a).

What are the limitations of land use scenarios?

Applying a common scenario framework (e.g., RCPs/SSPs) supports the comparison and integration of climate and land system scenarios, but a ‘climate-centric’ perspective can limit the capacity of these scenarios to account for a wider range of land-relevant drivers (Rosa et al. 2017). For example, in climate mitigation scenarios it is important to assess the impact of mitigation actions on the broader environment for example, biodiversity, ecosystem functioning, air quality, food security, desertification/degradation and water cycles (Rosa et al. 2017). This implies the need for a more encompassing and flexible approach to creating scenarios that considers other environmental aspects, not only as a part of impact assessment, but also during the process of creating the scenarios themselves.

A limited number of models can quantify global scale, land use change scenarios, and there is large variance in the outcomes of these models (Alexander et al. 2016a; Prestele et al. 2016). In some cases, there is greater variability between the models themselves than between the scenarios that they are quantifying, and these differences vary geographically (Prestele et al. 2016). These differences arise from variations in baseline datasets, thematic classes and modelling paradigms (Alexander et al. 2016a; Popp et al. 2016; Prestele et al. 2016). Model evaluation is critical in establishing confidence in the outcomes of modelled futures (Ahlstrom et al. 2012; Kelley et al. 2013). Some, but not all, land use models are evaluated against observational data and model evaluation is rarely reported. Hence, there is a need for more transparency in land use modelling, especially in evaluation and testing, as well as making model code available with complete sets of scenario outputs (e.g., Dietrich et al. 2018).

There is a small, but growing literature on quantitative pathways to achieve normative visions and their associated trade-offs (IPBES 2018a). Whilst the visions themselves may be clearly articulated, the societal choices, behaviours and transitions needed to attain them, are not. Better accounting for human behaviour and decision-making processes in global scale, land-use models would improve the capacity to quantify pathways to sustainable futures (Rounsevell et al. 2014; Arneth et al. 2014; Calvin and Bond-Lamberty 2018). It is, however, difficult to understand and represent human behaviour and social interaction processes at global scales. Decision-making in global models is commonly represented through economic processes (Arneth et al. 2014). Other important human processes for land systems including equity, fairness, land tenure and the role of institutions and governance, receive less attention, and this limits the use of global models to quantify transformative pathways, adaptation and mitigation (Arneth et al. 2014; Rounsevell et al. 2014; Wang et al. 2016). No model exists at present to represent complex human behaviours at the global scale, although the need has been highlighted (Rounsevell et al. 2014; Arneth et al. 2014; Robinson et al. 2017; Brown et al. 2017; Calvin and Bond-Lamberty 2018).

1
2 **1.2.2.3 Uncertainties in decision making**
3 Decision makers develop and implement policy in the face of many uncertainties (Rosenzweig and
4 Neofotis 2013; Anav et al. 2013; Ciais et al. 2013a; Stocker et al. 2013b; see Section 7.5). In context of
5 climate change, the term *deep uncertainty* is frequently used to denote situations in which either the
6 analysis of a situation is inconclusive, or parties to a decision cannot agree on a number of criteria that
7 would help to rank model results in terms of likelihood (e.g., Hallegatte and Mach 2016; Maier et al.
8 2016) (see Section 7.1, 7.5, and Supplementary Material Table SM. 1.2). However, existing uncertainty
9 does not support societal and political inaction.

10 The many ways of dealing with uncertainty in decision making can be summarised by two decision
11 approaches: (economic) cost-benefit analyses, and the precautionary approach. A typical variant of cost
12 benefit analysis is the minimisation of negative consequences. This approach needs reliable probability
13 estimates (Gleckler et al. 2016; Parker 2013) and tends to focus on the short-term. The precautionary
14 approach does not take account of probability estimates (cf. Raffensperger and Tickner 1999), but instead
15 focuses on avoiding the worst outcome (Gardiner 2006).

16 Between these two extremes, various decision approaches seek to address uncertainties in a more
17 reflective manner that avoids the limitations of cost-benefit analysis and the precautionary approach.
18 Climate-informed decision analysis combines various approaches to explore options and the
19 vulnerabilities and sensitivities of certain decisions. Such an approach includes stakeholder involvement
20 (e.g., elicitation methods), and can be combined with, for example, analysis of climate or land-use change
21 modelling (Hallegatte and Rentschler 2015; Luedeling and Shepherd 2016).

22 Flexibility is facilitated by political decisions that are not set in stone and can change over time (Walker et
23 al. 2013; Hallegatte and Rentschler 2015). Generally, within the research community that investigates
24 deep uncertainty a paradigm is emerging that requires to develop a strategic vision of the long- or mid-
25 term future, while committing to short-term actions and establishing a framework to guide future actions
26 including revisions and flexible adjustment of decisions (Haasnoot 2013; see Section 7.5).

27 **1.3 Response options to the key challenges**

28 A number of response options underpin solutions to the challenges arising from GHG emissions from
29 land, and the loss of productivity arising from degradation and desertification. These options are

1 discussed in Sections 2.5, 6.2 and rely on a) land management, b) value chain management and c) risk
 2 management (see Table 1.2). None of these response options are mutually exclusive, and it is their
 3 combination in a regionally, context-specific manner that is most likely to achieve co-benefits between
 4 climate change mitigation, adaptation and other environmental challenges in a cost- effective way
 5 (Griscom et al. 2017; Kok et al. 2018). Sustainable solutions affecting both demand and supply are
 6 expected to yield most co-benefits if these rely not only on the carbon footprint, but are extended to other
 7 vital ecosystems such as water, nutrients and biodiversity footprints (van Noordwijk and Brussaard 2014;
 8 Cremasch 2016). As an entry-point to the discussion in Chapter 6, we introduce here a selected number of
 9 examples that cut across climate change mitigation, food security, desertification, and degradation issues,
 10 including potential trade-offs and co-benefits.

11 **Table 1.2 Broad categorisation of response options into three main classes and eight sub-classes.** For
 12 illustration, the table includes examples of individual response options. A complete list and description is provided
 13 in Chapter 6.

| Response options based on land management | |
|--|---|
| <i>in agriculture</i> | Improved management of: cropland, grazing land, livestock; Agro-forestry; Avoidance of conversion of grassland to cropland; Integrated water management |
| <i>in forests</i> | Improved management of forests and forest restoration; Reduced deforestation and degradation; Afforestation |
| <i>of soils</i> | Increased soil organic carbon content; Reduced soil erosion; Reduced soil salinisation |
| <i>across all/other ecosystems</i> | Reduced landslides and natural hazards; Reduced pollution including acidification; Biodiversity conservation; Restoration and reduced conversion of peatlands |
| <i>specifically for carbon dioxide removal</i> | Enhanced weathering of minerals; Bioenergy and BECCS |
| Response options based on value chain management | |
| <i>through demand management</i> | Dietary change; Reduced post-harvest losses; Reduced food waste |
| <i>through supply management</i> | Sustainable sourcing; Improved energy use in food systems; Improved food processing and retailing |
| Response options based on risk management | |
| <i>risk management</i> | Risk sharing instruments; Use of local seeds; Disaster risk management |

14

15 1.3.1 Targeted decarbonisation relying on large land-area need

16 Most global future scenarios that aim to achieve global warming of 2°C or well below rely on bioenergy
 17 (BE; with or without carbon capture and storage, BECCS; see Cross-Chapter Box 7 in Chapter 6) or
 18 afforestation and reforestation (Cross-Chapter Box 2 in this Chapter)(de Coninck et al. 2018; Rogelj et al.
 19 2018b,a; Anderson and Peters 2016; Popp et al. 2016; Smith et al. 2016). In addition to the very large
 20 area requirements projected for 2050 or 2100, several other aspects of these scenarios have also been
 21 criticised. For instance, they simulate very rapid technological and societal uptake rates for the land-
 22 related mitigation measures, when compared with historical observations (Turner et al. 2018; Brown et al.
 23 2019; Vaughan and Gough 2016). Furthermore, confidence in the projected bioenergy or BECCS net
 24 carbon uptake potential is *low*, because of many diverging assumptions. This includes assumptions about
 25 bioenergy crop yields, the possibly large energy demand for CCS, which diminishes the net-GHG-saving

1 of bioenergy systems, or the incomplete accounting for ecosystem processes and of the cumulative
2 carbon-loss arising from natural vegetation clearance for bioenergy crops or bioenergy forests and
3 subsequent harvest regimes (Anderson and Peters 2016; Bentsen 2017; Searchinger et al. 2017; Bayer et
4 al. 2017; Fuchs et al. 2017; Pingoud et al. 2018; Schlesinger 2018). Bioenergy provision under politically
5 unstable conditions may also be a problem (Erb et al. 2012; Searle and Malins 2015).

6 Large-scale bioenergy plantations and forests may compete for the same land area (Harper et al. 2018).
7 Both potentially have adverse side effects on biodiversity and ecosystem services, as well as socio-
8 economic trade-offs such as higher food prices due to land area competition (Shi et al. 2013; Bárcena et al.
9 2014; Fernandez-Martinez et al. 2014; Searchinger et al. 2015; Bonsch et al. 2016; Creutzig et al. 2015;
10 Kreidenweis et al. 2016; Santangeli et al. 2016; Williamson 2016; Graham et al. 2017; Krause et al. 2017;
11 Hasegawa et al. 2018; Humpenoeder et al. 2018). Although forest-based mitigation could have co-
12 benefits for biodiversity and many ecosystem services, this depends on the type of forest planted and the
13 vegetation cover it replaces (Popp et al. 2014; Searchinger et al. 2015) (see also Cross-Chapter Box 2 in
14 this Chapter).

15 There is *high confidence* that scenarios with large land requirements for climate change mitigation may
16 not achieve sustainable development goals, such as no poverty, zero hunger and life on land, if
17 competition for land and the need for agricultural intensification are greatly enhanced (Creutzig et al.
18 2016; Dooley and Kartha 2018; Hasegawa et al. 2015; Hof et al. 2018; Roy et al. 2018; Santangeli et al.
19 2016; Boysen et al. 2017; Henry et al. 2018; Kreidenweis et al. 2016; UN 2015). This does not mean that
20 smaller-scale land-based climate mitigation can have positive outcomes for then achieving these goals
21 (see e.g., Sections 6.2, 4.5, cross chapter box 7 in Chapter 6).

Cross-Chapter Box 2: Implications of large-scale conversion from non-forest to forest land

Baldur Janz (Germany), Almut Arneth (Germany), Francesco Cherubini (Norway/Italy), Edouard Davin (Switzerland/France), Aziz Elbehri (Morocco), Kaoru Kitajima (Japan), Werner Kurz (Canada)

Efforts to increase forest area

While deforestation continues in many world regions, especially in the tropics, large expansion of mostly managed forest area has taken place in some countries. In the IPCC context, reforestation (conversion to forest of land that previously contained forests but has been converted to some other use) is distinguished from afforestation (conversion to forest of land that historically has not contained forests (see glossary)). Past expansion of managed forest area occurred in many world-regions for a variety of reasons, from meeting needs for wood fuel or timber (Vadell et al. 2016; Joshi et al. 2011; Zaloumis and Bond 2015; Payn et al. 2015; Shoyama 2008; Miyamoto et al. 2011) to restoration-driven efforts, with the aim of enhancing ecological function (Filoso et al. 2017; Salvati and Carlucci 2014; Ogle et al. 2018; Crouzeilles et al. 2016; FAO 2016)(see Section 3.7, 4.9).

In many regions, net forest area increase includes deforestation (often of native forests) alongside increasing forest area (often managed forest, but also more natural forest restoration efforts; (Heilmayr et al. 2016; Scheidel and Work 2018; Hua et al. 2018; Crouzeilles et al. 2016; Chazdon et al. 2016b). China and India have seen the largest net forest area increase, aiming to alleviate soil erosion, desertification and overgrazing (Ahrends et al. 2017; Cao et al. 2016; Deng et al. 2015; Chen et al. 2019)(see Section 3.7, 4.9) but uncertainties in exact forest area changes remain large, mostly due to differences in methodology and forest classification (FAO 2015a; Song et al. 2018; Hansen et al. 2013; MacDicken et al. 2015)(Section 1.1.2).

What are the implications for ecosystems?

1) Implications for biogeochemical and biophysical processes

There is *robust evidence* and *medium agreement* that whilst forest area expansion increases ecosystem carbon storage, the magnitude of the increased stock depends on the type and length of former land-use, forest type planted, and climatic regions (Bárcena et al. 2014; Poeplau et al. 2011; Shi et al. 2013; Li et al. 2012)(see Section 4.3). While, reforestation of former croplands increases net ecosystem carbon storage (Bernal et al. 2018; Lamb 2018), afforestation on native grassland results in reduction of soil carbon stocks, which can reduce or negate the net carbon benefits which are dominated by increases in biomass, dead wood and litter carbon pools (Veldman et al. 2015, 2017).

Forest vs. non-forest lands differ in land surface reflectiveness of short-wave radiation and evapotranspiration (Anderson et al. 2011; Perugini et al. 2017)(see Section 2.4). Evapotranspiration from forests during the growing season regionally cools the land surface and enhances cloud cover that reduces short wave radiation reaching the land, an impact that is especially pronounced in the tropics. However, dark evergreen conifer-dominated forests have low surface reflectance, and tend to cause warming of the near surface atmosphere compared to non-forest land, especially when snow cover is present such as in boreal regions (Duveiller et al. 2018; Alkama and Cescatti 2016; Perugini et al. 2017)(*medium evidence, high agreement*).

2) Implications for water balance

Evapotranspiration by forests reduces surface runoff and erosion of soil and nutrients (Salvati et al.

2014). Planting of fast-growing species in semi-arid regions or replacing natural grasslands with forest plantations can divert soil water resources to evapotranspiration from groundwater recharge (Silveira et al. 2016; Zheng et al. 2016; Cao et al. 2016). Multiple cases are reported from China where afforestation programs, some with irrigation, without having tailored to local precipitation conditions, resulted in water shortages and tree mortality (Cao et al. 2016; Yang et al. 2014; Li et al. 2014; Feng et al. 2016). Water shortages may create long-term water conflicts (Zheng et al. 2016). However, reforestation (in particular for restoration) is also associated with improved water filtration, groundwater recharge (Ellison et al. 2017) and can reduce risk of soil erosion, flooding, and associated disasters (Lee et al. 2018; see Section 4.9).

3) Implications for biodiversity

Impacts of forest area expansion on biodiversity depend mostly on the vegetation cover that is replaced: afforestation on natural non-tree dominated ecosystems can have negative impacts on biodiversity (Abreu et al. 2017; Griffith et al. 2017; Veldman et al. 2015; Parr et al. 2014; Wilson et al. 2017; Hua et al. 2016)(see also IPCC 1.5° report (2018). Reforestation with monocultures of fast growing, non-native trees has little benefit to biodiversity (Shimamoto et al. 2018; Hua et al. 2016). There are also concerns regarding the impacts of some commonly used plantation species (e.g., *Acacia* and *Pinus* species) to become invasive (Padmanaba and Corlett 2014; Cunningham et al. 2015b).

Reforestation with mixes of native species, especially in areas that retain fragments of native forest, can support ecosystem-services and biodiversity recovery, with positive social and environmental co-benefits (Cunningham et al. 2015a; Dendy et al. 2015; Chaudhary and Kastner 2016; Huang et al. 2018; Locatelli et al. 2015b)(see Section 4.5). Even though species diversity in re-growing forests is typically lower than in primary forests, planting native or mixed species can have positive effects on biodiversity (Brockerhoff et al. 2013; Pawson et al. 2013; Thompson et al. 2014). Reforestation has been shown to improve links among existing remnant forest patches, increasing species movement, and fostering gene flow between otherwise isolated populations (Gilbert-Norton et al. 2010; Barlow et al. 2007; Lindenmayer and Hobbs 2004).

4) Implications for other ecosystem services and societies

Forest area expansion could benefit recreation and health, preservation of cultural heritage and local values and knowledge, livelihood support (via reduced resource conflicts, restoration of local resources). These social benefits could be most successfully achieved if local communities' concerns are considered (Le et al. 2012). However, these co-benefits have rarely been assessed due to a lack of suitable frameworks and evaluation tools (Baral et al. 2016).

Industrial forest management can be in conflict with needs of forest-dependent people and community-based forest management over access to natural resources (Gerber 2011; Baral et al. 2016) and/or loss of customary rights over land use (Malkamäki et al. 2018; Cotula et al. 2014). A common result is out-migration from rural areas and diminishing local uses of ecosystems (Gerber 2011). Policies promoting large-scale tree plantations gain if these are reappraised in view of potential co-benefits with several ecosystem services and local societies (Bull et al. 2006; Le et al. 2012).

Scenarios of forest-area expansion for land-based climate change mitigation

Conversion of non-forest to forest land has been discussed as a relatively cost-effective climate change mitigation option when compared to options in the energy and transport sectors (*medium evidence, medium agreement*) (de Coninck et al. 2018; Griscom et al. 2017; Fuss et al. 2018), and can have co-

benefits with adaptation.

Sequestration of CO₂ from the atmosphere through forest area expansion has become a fundamental part of stringent climate change mitigation scenarios (Rogelj et al. 2018a; Fuss et al. 2018)(see e.g., Sections 2.5, 4.5, 6.2). The estimated mitigation potential ranges from about 0.5 to 10 Gt CO₂yr⁻¹ (*robust evidence, medium agreement*), and depends on assumptions regarding available land and forest carbon uptake potential (Houghton 2013; Houghton and Nassikas 2017; Griscom et al. 2017; Lenton 2014; Fuss et al. 2018; Smith 2016) (see Section 2.5.1). In climate change mitigation scenarios, typically, no differentiation is made between reforestation and afforestation despite different overall environmental impacts between these two measures. Likewise, biodiversity conservation, impacts on water balances, other ecosystem services, or land-ownership as constraints when simulating forest area expansion (see Cross-Chapter Box 1 in this Chapter) tend not to be included as constraints when simulating forest area expansion.

Projected forest area increases, relative to today's forest area, range from approximately 25% in 2050 and increase to nearly 50% by 2100 (Rogelj et al. 2018a; Kreidenweis et al. 2016; Humpenoder et al. 2014). Potential adverse side-effects of such large-scale measures, especially for low-income countries, could be increasing food prices from the increased competition for land (Kreidenweis et al. 2016; Hasegawa et al. 2015, 2018; Boysen et al. 2017)(see Section 5.5). Forests also emit large amounts of biogenic volatile compounds that under some conditions contribute to the formation of atmospherically short-lived climate forcing compounds, which are also detrimental to health (Ashworth et al. 2013; Harrison et al. 2013). Recent analyses argued for an upper limit of about 5 million km² of land globally available for climate change mitigation through reforestation, mostly in the tropics (Houghton 2013) – with potential regional co-benefits.

Since forest growth competes for land with bioenergy crops (Harper et al. 2018)(Cross-Chapter Box 7: Bioenergy and BECCS, Chapter 6), global area estimates need to be assessed in light of alternative mitigation measures at a given location. In all forest-based mitigation efforts, the sequestration potential will eventually saturate unless the area keeps expanding, or harvested wood is either used for long-term storage products or for carbon capture and storage (Fuss et al. 2018; Houghton et al. 2015)(see Section 2.5.1). Considerable uncertainty in forest carbon uptake estimates is further introduced by potential forest losses from fire or pest outbreaks (Allen et al. 2010; Anderegg et al. 2015)(Cross-Chapter Box 3: Fire and climate change, Chapter 2). And like all land-based mitigation measures, benefits may be diminished by land-use displacement, through trade of land-based products, especially in poor countries that experience forest loss (e.g., Africa) (Bhojvaid et al. 2016; Jadin et al. 2016).

Conclusion

Reforestation is a mitigation measure with potential co-benefits for conservation and adaptation, including biodiversity habitat, air and water filtration, flood control, enhanced soil fertility and reversal of land degradation. Potential adverse side-effects of forest area expansion depend largely on the state of the land it displaces as well as tree species selections. Active governance and planning contribute to maximising co-benefits while minimising adverse side-effects (Laestadius et al. 2011; Dinerstein et al. 2015; Veldman et al. 2017)(see Section 4.8 and Chapter 7). At large spatial scales, forest expansion is expected to lead to increased competition for land, with potentially undesirable impacts on food prices, biodiversity, non-forest ecosystems and water availability (Bryan and Crossman 2013; Boysen et al. 2017; Kreidenweis et al. 2016; Egginton et al. 2014; Cao et al. 2016; Locatelli et al. 2015a; Smith et al.

2013)

1

2 **1.3.2 Land Management**

3 ***1.3.2.1 Agricultural, forest and soil management***

4 Sustainable land management (SLM) describes “*the stewardship and use of land resources, including*
5 *soils, water, animals and plants, to meet changing human needs while simultaneously assuring the long-*
6 *term productive potential of these resources and the maintenance of their environmental functions*”
7 (Alemu 2016, Altieri and Nicholls 2017)(see e.g., Section 4.1.5), and includes ecological, technological
8 and governance aspects.

9 The choice of SLM strategy is a function of regional context and land use types, with *high agreement* on
10 (a combination of) choices such as agroecology (including agroforestry), conservation agriculture and
11 forestry practices, crop and forest species diversity, appropriate crop and forest rotations, organic farming,
12 integrated pest management, the preservation and protection of pollination services, rain water harvesting,
13 range and pasture management, and precision agriculture systems (Stockmann et al. 2013; Ebert, 2014;
14 Schulte et al. 2014; Zhang et al. 2015; Sunil and Pandravada 2015; Poepflau and Don 2015; Agus et al.
15 2015; Keenan 2015; MacDicken et al. 2015; Abberton et al. 2016). Conservation agriculture and forestry
16 uses management practises with minimal soil disturbance such as no tillage or minimum tillage,
17 permanent soil cover with mulch combined with rotations to ensure a permanent soil surface, or rapid
18 regeneration of forest following harvest (Hobbs et al. 2008; Friedrich et al. 2012). Vegetation and soils in
19 forests and woodland ecosystems play a crucial role in regulating critical ecosystem processes, therefore
20 reduced deforestation together with sustainable forest management are integral to SLM (FAO 2015b; see
21 Section 4.8). In some circumstances, increased demand for forest products can also lead to increased
22 management of carbon storage in forests (Favero and Mendelsohn 2014). Precision agriculture is
23 characterised by a “management system that is information and technology based, is site specific and uses
24 one or more of the following sources of data: soils, crops, nutrients, pests, moisture, or yield, for optimum
25 profitability, sustainability, and protection of the environment” (USDA 2007)(see also Cross-Chapter Box
26 6: Agricultural intensification, Chapter 5). The management of protected areas that reduce deforestation
27 also plays an important role in climate change mitigation and adaptation while delivering numerous
28 ecosystem services and sustainable development benefits (Bebber and Butt 2017). Similarly, when
29 managed in an integrated and sustainable way, peatlands are also known to provide numerous ecosystem
30 services, as well as socio-economic and mitigation and adaptation benefits (Ziadat et al. 2018).

31 Biochar is an organic compound used as soil amendment and is believed to be potentially an important
32 global resource for mitigation. Enhancing the carbon content of soil and/or use of biochar (see Chapter 4)
33 have become increasingly important as a climate change mitigation option with possibly large co-benefits
34 for other ecosystem services. Enhancing soil carbon storage and the addition of biochar can be practised
35 with limited competition for land, provided no productivity/yield loss and abundant unused biomass, but
36 evidence is limited and impacts of large scale application of biochar on the full GHG balance of soils, or
37 human health are yet to be explored (Gurwick et al. 2013; Lorenz and Lal 2014; Smith 2016).

38 **1.3.3 Value chain management**

39 ***1.3.3.1 Supply management***

40 **Food losses from harvest to retailer.** Approximately one third of losses and waste in the food system
41 occurs between crop production and food consumption, increasing substantially if losses in livestock

1 production and overeating are included (Gustavsson et al. 2011; Alexander et al. 2017). This includes on-
2 farm losses, farm to retailer losses, as well retailer and consumer losses (see Section 1.3.3.2).

3 Post-harvest food loss on farm and from farm to retailer is a widespread problem, especially in
4 developing countries (Xue et al. 2017), but are challenging to quantify. For instance, averaged for eastern
5 and southern Africa an estimated 10–17% of annual grain production is lost (Zorya et al. 2011). Across
6 84 countries and different time periods, annual median losses in the supply chain before retailing were
7 estimated at about 28 kg per capita for cereals or about 12 kg per capita for eggs and dairy products (Xue
8 et al. 2017). For the year 2013, losses prior to the reaching retailers were estimated at 20% (dry weight) of
9 the production amount (22% wet weight) (Gustavsson et al. 2011; Alexander et al. 2017). While losses of
10 food cannot be realistically reduced to zero, advancing harvesting technologies (Bradford et al. 2018;
11 Affognon et al. 2015), storage capacity (Chegere 2018) and efficient transportation could all contribute to
12 reducing these losses with co-benefits for food availability, the land area needed for food production and
13 related GHG emissions.

14 **Stability of food supply, transport and distribution.** Increased climate variability enhances fluctuations
15 in world food supply and price variability (Warren 2014; Challinor et al. 2015; Elbehri et al. 2017). “Food
16 price shocks” need to be understood regarding their transmission across sectors and borders and impacts
17 on poor and food insecure populations, including urban poor subject to food deserts and inadequate food
18 accessibility (Widener et al. 2017; Lehmann et al. 2013; LE 2016; FAO 2015b). Trade can play an
19 important stabilising role in food supply, especially for regions with agro-ecological limits to production,
20 including water scarce regions, as well as regions that experience short term production variability due to
21 climate, conflicts or other economic shocks (Gilmont 2015; Marchand et al. 2016). Food trade can either
22 increase or reduce the overall environmental impacts of agriculture (Kastner et al. 2014). Embedded in
23 trade are virtual transfers of water, land area, productivity, ecosystem services, biodiversity, or nutrients
24 (Marques et al. 2019; Wiedmann and Lenzen 2018; Chaudhary and Kastner 2016) with either positive or
25 negative implications (Chen et al. 2018; Yu et al. 2013). Detrimental consequences in countries in which
26 trade dependency may accentuate the risk of food shortages from foreign production shocks could be
27 reduced by increasing domestic reserves or importing food from a diversity of suppliers (Gilmont 2015;
28 Marchand et al. 2016).

29 Climate mitigation policies could create new trade opportunities (e.g., biomass) (Favero and Massetti
30 2014) or alter existing trade patterns. The transportation GHG-footprints of supply chains may be causing
31 a differentiation between short and long supply chains (Schmidt et al. 2017) that may be influenced by
32 both economics and policy measures (see Section 5.4). In the absence of sustainable practices and when
33 the ecological footprint is not valued through the market system, trade can also exacerbate resource
34 exploitation and environmental leakages, thus weakening trade mitigation contributions (Dalin and
35 Rodríguez-Iturbe 2016; Mosnier et al. 2014; Elbehri et al. 2017). Ensuring stable food supply while
36 pursuing climate mitigation and adaptation will benefit from evolving trade rules and policies that allow
37 internalisation of the cost of carbon (and costs of other vital resources such as water, nutrients). Likewise,
38 future climate change mitigation policies would gain from measures designed to internalise the
39 environmental costs of resources and the benefits of ecosystem services (Elbehri et al. 2017; Brown et
40 al., 2007).

41 **1.3.3.2 Demand management**

42 **Dietary change.** Demand-side solutions to climate mitigation are an essential complement to supply-side,
43 technology and productivity driven solutions (Creutzig et al. 2016; Bajželj et al. 2014; Erb et al. 2016b;
44 Creutzig et al. 2018)(see Sections 5.5.1, 5.5.2)(*high confidence*). The environmental impacts of the

1 animal-rich “western diets” are being examined critically in the scientific literature (Hallström et al. 2015;
2 Alexander et al. 2016b; Alexander et al. 2015; Tilman and Clark 2014; Aleksandrowicz et al. 2016; Poore
3 and Nemecek 2018)(see Section 5.4.6). For example, if the average diet of each country were consumed
4 globally, the agricultural land area needed to supply these diets would vary 14-fold, due to country
5 differences in ruminant protein and calorific intake (-55% to +178% compared to existing cropland
6 areas). Given the important role enteric fermentation plays in methane (CH₄) emissions, a number of
7 studies have examined the implications of lower animal diets (Swain et al. 2018; Rööös et al. 2017; Rao et
8 al. 2018). Reduction of animal protein intake has been estimated to reduce global green water (from
9 precipitation) use by 11% and blue water (from rivers, lakes, groundwater) use by 6% (Jalava et al. 2014).
10 By avoiding meat from producers with above-median GHG emissions and halving animal-product intake,
11 consumption change could free-up 21 million km² of agricultural land and reduce GHG emissions by
12 nearly 5 Gt CO₂-eq yr⁻¹ or up to 10.4 Gt CO₂-eq yr⁻¹ when vegetation carbon uptake is considered on the
13 previously agricultural land (Poore and Nemecek 2018, 2019).

14 Diets can be location and community specific, are rooted in culture and traditions while responding to
15 changing lifestyles driven for instance by urbanisation and changing income. Changing dietary and
16 consumption habits would require a combination of non-price (government procurement, regulations,
17 education and awareness raising) and price (Juhl and Jensen 2014) incentives to induce consumer
18 behavioural change with potential synergies between climate, health and equity (addressing growing
19 global nutrition imbalances that emerge as undernutrition, malnutrition, and obesity) (FAO 2018b).

20 **Reduced waste and losses in the food demand system.** Global averaged per capita food waste and loss
21 (FWL) have increased by 44% between 1961 and 2011 (Porter et al. 2016) and are now around 25–30%
22 of global food produced (Kummu et al. 2012)(Alexander et al. 2017). Food waste occurs at all stages of
23 the food supply chain from the household to the marketplace (Parfitt et al. 2010) and is found to be larger
24 at household than at supply chain levels. A meta-analysis of 55 studies showed that the highest share of
25 food waste was at the consumer stage (43.9% of total) with waste increasing with per capita GDP for high
26 income countries until a plateau at about 100 kg cap⁻¹ yr⁻¹ (around 16% of food consumption) above
27 about 70 000 USD cap⁻¹ (van der Werf and Gilliland 2017; Xue et al. 2017). Food loss from supply chains
28 tends to be more prevalent in less developed countries where inadequate technologies, limited
29 infrastructure, and imperfect markets combine to raise the share of the food production lost before use.

30 There are several causes behind food waste including economics (cheap food), food policies (subsidies)
31 as well as individual behaviour (Schanes et al. 2018). Household level food waste arises from overeating
32 or overbuying (Thyberg and Tonjes 2016). Globally, overconsumption was found to waste 9–10% of
33 food bought (Alexander et al. 2017).

34 Solutions to FWL thus need to address technical and economic aspects. Such solutions would benefit
35 from more accurate data on the loss-source, -magnitude and -causes along the food supply chain. In the
36 long run, internalising the cost of food waste into the product price would more likely induce a shift in
37 consumer behaviour towards less waste and more nutritious, or alternative, food intake (FAO 2018b).
38 Reducing FWL would bring a range of benefits for health, reducing pressures on land, water and
39 nutrients, lowering emissions and safeguarding food security. Reducing food waste by 50% would
40 generate net emissions reductions in the range of 20 to 30% of total food-sourced GHGs (Bajželj et al.
41 2014). The SDG 12 (“Ensure sustainable consumption and production patterns”) calls for per capita
42 global food waste to be reduced by one half at the retail and consumer level, and reducing food losses
43 along production and supply chains by 2030.

1 **1.3.4 Risk management**

2 Risk management refers to plans, actions, strategies or policies to reduce the likelihood and/or magnitude
3 of adverse potential consequences, based on assessed or perceived risks' Insurance and early warning
4 systems are examples of risk management, but risk can also be reduced (or resilience enhanced) through a
5 broad set of options ranging from seed sovereignty, livelihood diversification, to reducing land loss
6 through urban sprawl. Early warning systems support farmer decision making on management strategies
7 (see Section 1.2) and are a good example of an adaptation measure with mitigation co-benefits such as
8 reducing carbon losses (see Section 1.3.6). Primarily designed to avoid yield losses, early warning
9 systems also support fire management strategies in forest ecosystems, which prevents financial as well as
10 carbon losses (de Groot et al. 2015). Given that over recent decades on average around 10% of cereal
11 production was lost through extreme weather events (Lesk et al. 2016), where available and affordable,
12 insurance can buffer farmers and foresters against the financial losses incurred through such weather and
13 other (fire, pests) extremes (Falco et al. 2014)(see Section 7.2, 7.4). Decisions to take up insurance are
14 influenced by a range of factors such as the removal of subsidies or targeted education (Falco et al. 2014).
15 Enhancing access and affordability of insurance in low-income countries is a specific objective of the
16 UNFCCC (Linnerooth-Bayer and Mechler 2006). A global mitigation co-benefit of insurance schemes
17 may also include incentives for future risk reduction (Surminski and Oramas-Dorta 2014).

18 **1.3.5 Economics of land-based mitigation pathways: Costs versus benefits of early action** 19 **under uncertainty**

20 The overarching societal costs associated with GHG emissions and the potential implications of
21 mitigation activities can be measured by various metrics (cost-benefit analysis, cost effectiveness
22 analysis) at different scales (project, technology, sector or the economy) (IPCC 2018; section 1.4). The
23 Social Cost of Carbon (SCC), measures the total net damages of an extra metric ton of CO₂ emissions due
24 to the associated climate change (Nordhaus 2014; Pizer et al. 2014). Both negative and positive impacts
25 are monetised and discounted to arrive at the net value of consumption loss. As the SCC depends on
26 discount rate assumptions and value judgements (e.g., relative weight given to current vs. future
27 generations), it is not a straightforward policy tool to compare alternative options. At the sectoral level,
28 marginal abatement cost curves (MACCs) are widely used for the assessment of costs related to GHG
29 emissions reduction. MACCs measure the cost of reducing one more GHG unit and are either expert-
30 based or model-derived and offer a range of approaches and assumptions on discount rates or available
31 abatement technologies (Kesicki 2013). In land-based sectors, Gillingham and Stock (2018) reported
32 short term static abatement costs for afforestation of between 1 and 10 USD 2017/tCO₂, soil management
33 at 57 and livestock management at 71 USD 2017/tCO₂. MACCs are more reliable when used to rank
34 alternative options compared to a baseline (or business as usual) rather than offering absolute numerical
35 measures (Huang et al. 2016). The economics of land-based mitigation options encompass also the "costs
36 of inaction" that arise either from the economic damages due to continued accumulation of GHGs in the
37 atmosphere and from the diminution in value of ecosystem services or the cost of their restoration where
38 feasible (Rodriguez-Labajos 2013; Ricke et al. 2018). Overall, it remains challenging to estimate the costs
39 of alternative mitigation options owing to the context- and scale specific interplay between multiple
40 drivers (technological, economic, and socio-cultural) and enabling policies and institutions (IPCC
41 2018)(section 1.4).

42 The costs associated with mitigation (both project-linked such as capital costs or land rental rates or
43 sometimes social costs) generally increase with stringent mitigation targets and over time. Sources of
44 uncertainty include the future availability, cost and performance of technologies (Rosen and Guenther

1 2015; Chen et al. 2016) or lags in decision making, which have been demonstrated by the uptake of land
2 use and land utilisation policies (Alexander et al. 2013; Hull et al. 2015; Brown et al. 2018b). There is
3 growing evidence of significant mitigation gains through conservation, restoration and improved land
4 management practices (Griscom et al. 2017; Kindermann et al. 2008; Golub et al. 2013; Favero et al.
5 2017)(see Chapter 4 and Chapter 6), but the mitigation cost efficiency can vary according to region and
6 specific ecosystem (Albanito et al. 2016). Recent model developments that treat process-based, human-
7 environment interactions have recognised feedbacks that reinforce or dampen the original stimulus for
8 land use change (Robinson et al. 2017; Walters and Scholes 2017). For instance, land mitigation
9 interventions that rely on large-scale, land use change (i.e., afforestation) would need to account for the
10 rebound effect (which dampens initial impacts due to feedbacks) in which raising land prices also raises
11 the cost of land-based mitigation (Vivanco et al. 2016). Although there are few direct estimates, indirect
12 assessments strongly point to much higher costs if action is delayed or limited in scope (*medium*
13 *confidence*). Quicker response options are also needed to avoid loss of high-carbon ecosystems and other
14 vital ecosystem services that provide multiple services that are difficult to replace (peatlands, wetlands,
15 mangroves, forests) (Yirdaw et al. 2017; Pedrozo-Acuña et al. 2015). Delayed action would raise relative
16 costs in the future or could make response options less feasible (Goldstein et al. 2019; Butler et al.
17 2014)(*medium confidence*).

18 **1.3.6 Adaptation measures and scope for co-benefits with mitigation**

19 Adaptation and mitigation have generally been treated as two separate discourses, both in policy and
20 practice with mitigation addressing cause and adaptation dealing with the consequences of climate change
21 (Hennessey et al. 2017). While adaptation (e.g., reducing flood risks) and mitigation (e.g., reducing non-
22 CO₂ emissions from agriculture) may have different objectives and operate at different scales, they can
23 also generate joint outcomes (Locatelli et al. 2015b) with adaptation generating mitigation co-benefits.
24 Seeking to integrate strategies for achieving adaptation and mitigation goals is attractive in order to
25 reduce competition for limited resources and trade-offs (Lobell et al. 2013; Berry et al. 2015; Kongsager
26 and Corbera 2015). Moreover, determinants that can foster adaptation and mitigation practices are
27 similar. These tend to include available technology and resources, and credible information for policy
28 makers to act on (Yohe 2001).

29 Four sets of mitigation-adaptation interrelationships can be distinguished: 1) mitigation actions that can
30 result in adaptation benefits; 2) adaptation actions that have mitigation benefits; 3) processes that have
31 implications for both adaptation and mitigation; 4) strategies and policy processes that seek to promote an
32 integrated set of responses for both adaptation and mitigation (Klein et al. 2007). A high level of adaptive
33 capacity is a key ingredient to developing successful mitigation policy. Implementing mitigation action
34 can result in increasing resilience especially if it is able to reduce risks. Yet, mitigation and adaptation
35 objectives, scale of implementation, sector and even metrics to identify impacts tend to differ (Ayers and
36 Huq 2009), and institutional setting, often does not enable an environment where synergies are sought
37 (Kongsager et al. 2016). Trade-offs between adaptation and mitigation exist as well and need to be
38 understood (and avoided) to establish win-win situations (Porter et al. 2014; Kongsager et al. 2016).

39 Forestry and agriculture offer a wide range of lessons for the integration of adaptation and mitigation
40 actions given the vulnerability of forest ecosystems or cropland to climate variability and change (Keenan
41 2015; Gaba et al. 2015)(see Section 5.6, 4.8). Increasing adaptive capacity in forested areas has the
42 potential to prevent deforestation and forest degradation (Locatelli et al. 2011). Reforestation projects, if
43 well managed, can increase community economic opportunities that encourage conservation (Nelson and
44 de Jong 2003), build capacity through training of farmers and installation of multifunctional plantations

1 with income generation (Reyer et al. 2009), strengthen local institutions (Locatelli et al. 2015a) and
2 increase cash-flow to local forest stakeholders from foreign donors (West 2016). A forest plantation that
3 sequesters carbon for mitigation can also reduce water availability to downstream populations and
4 heighten their vulnerability to drought. Inversely, not recognising mitigation in adaptation projects may
5 yield adaptation measures that increase greenhouse gas emissions, a prime example of ‘maladaptation’.
6 Analogously, ‘mal-mitigation’ would result in reducing greenhouse gas emissions, but increasing
7 vulnerability (Barnett and O’Neill 2010; Porter et al. 2014). For instance, the cost of pursuing large scale
8 adaptation and mitigation projects has been associated with higher failure risks, onerous transactions costs
9 and the complexity of managing big projects (Swart and Raes 2007).

10 Adaptation encompasses both biophysical and socio-economic vulnerability and underlying causes
11 (informational, capacity, financial, institutional, and technological; Huq et al. 2014) and it is increasingly
12 linked to resilience and to broader development goals (Huq et al. 2014). Adaptation measures can
13 increase performance of mitigation projects under climate change and legitimise mitigation measures
14 through the more immediately felt effects of adaptation (Locatelli et al. 2011; Campbell et al. 2014;
15 Locatelli et al. 2015b). Effective climate policy integration in the land sector is expected to gain from 1)
16 internal policy coherence between adaptation and mitigation objectives, 2) external climate coherence
17 between climate change and development objectives, 3) policy integration that favours vertical
18 governance structures to foster effective mainstreaming of climate change into sectoral policies, and 4)
19 horizontal policy integration through overarching governance structures to enable cross-sectoral co-
20 ordination (see Sections 1.4, 7.4).

21 **1.4 Enabling the response**

22 Climate change and sustainable development are challenges to society that require action at local,
23 national, transboundary and global scales. Different time-perspectives are also important in decision
24 making, ranging from immediate actions to long-term planning and investment. Acknowledging the
25 systemic link between food production and consumption, and land-resources more broadly is expected to
26 enhance the success of actions (Bazilian et al. 2011; Hussey and Pittock 2012). Because of the complexity
27 of challenges and the diversity of actors involved in addressing these challenges, decision making would
28 benefit from a portfolio of policy instruments. Decision making would also be facilitated by overcoming
29 barriers such as inadequate education and funding mechanisms, as well as integrating international
30 decisions into all relevant (sub)national sectoral policies (see Section 7.4).

31 ‘Nexus thinking’ emerged as an alternative to the sector-specific governance of natural resource use to
32 achieve global securities of water (D’Odorico et al. 2018), food and energy (Hoff 2011; Allan et al.
33 2015), and also to address biodiversity concerns (Fischer et al. 2017). Yet, there is no agreed definition of
34 “nexus” nor a uniform framework to approach the concept, which may be land-focused (Howells et al.
35 2013), water-focused (Hoff 2011) or food-centred (Ringler and Lawford 2013; Biggs et al. 2015).
36 Significant barriers remain to establish nexus approaches as part of a wider repertoire of responses to
37 global environmental change, including challenges to cross-disciplinary collaboration, complexity,
38 political economy and the incompatibility of current institutional structures (Hayley et al. 2015; Wichelns
39 2017)(see Section 7.5.6, 7.6.2).

40 **1.4.1 Governance to enable the response**

41 Governance includes the processes, structures, rules and traditions applied by formal and informal actors
42 including governments, markets, organisations, and their interactions with people. Land governance
43 actors include those affecting policies and markets, and those directly changing land use (Hersperger et al.

1 2010). The former includes governments and administrative entities, large companies investing in land,
2 non-governmental institutions and international institutions. It also includes UN agencies that are working
3 at the interface between climate change and land management, such as the FAO and the World Food
4 Programme that have *inter alia* worked on advancing knowledge to support food security through the
5 improvement of techniques and strategies for more resilient farm systems. Farmers and foresters directly
6 act on land (actors in proximate causes) (Hersperger et al. 2010)(see also Chapter 7.).

7 Policy design and formulation has often been strongly sectoral. For example, agricultural policy might be
8 concerned with food security, but have little concern for environmental protection or human health. As
9 food, energy and water security and the conservation of biodiversity rank highly on the Agenda 2030 for
10 Sustainable Development, the promotion of synergies between and across sectoral policies is important
11 (IPBES 2018a). This can also reduce the risks of anthropogenic climate forcing through mitigation, and
12 bring greater collaboration between scientists, policy makers, the private sector and land managers in
13 adapting to climate change (FAO 2015a). Polycentric governance (see Section 7.6) has emerged as an
14 appropriate way of handling resource management problems, in which the decision- making centers take
15 account of one another in competitive and cooperative relationships and have recourse to conflict
16 resolution mechanisms (Carlisle and Gruby 2017). Polycentric governance is also multi-scale and allows
17 the interaction between actors at different levels (local, regional, national, and global) in managing
18 common pool resources such as forests or aquifers.

19 Implementation of systemic, nexus approaches has been achieved through socio-ecological systems (SES)
20 frameworks that emerged from studies of how institutions affect human incentives, actions and outcomes
21 (Ostrom and Cox 2010). Recognition of the importance of SES laid the basis for alternative formulations
22 to tackle the sustainable management of land resources focusing specifically on institutional and
23 governance outcomes (Lebel et al. 2006; Bodin 2017). The SES approach also addresses the multiple
24 scales in which the social and ecological dimensions interact (Veldkamp et al. 2011; Myers et al. 2016;
25 Azizi et al. 2017) (see Section 6.1).

26 Adaptation or resilience pathways within the SES frameworks require several attributes, including
27 indigenous and local knowledge (ILK) and trust building for deliberative decision making and effective
28 collective action, polycentric and multi-layered institutions and responsible authorities that pursue just
29 distributions of benefits to enhance the adaptive capacity of vulnerable groups and communities (Lebel et
30 al. 2006). The nature, source, and mode of knowledge generation are critical to ensure that sustainable
31 solutions are community-owned and fully integrated within the local context (Mistry and Berardi 2016;
32 Schneider and Buser 2018). Integrating ILK with scientific information is a prerequisite for such
33 community-owned solutions (see Cross-Chapter Box 13: ILK, Chapter 7). ILK is context-specific,
34 transmitted orally or through imitation and demonstration, adaptive to changing environments,
35 collectivised through a shared social memory (Mistry and Berardi 2016). ILK is also holistic since
36 indigenous people do not seek solutions aimed at adapting to climate change alone, but instead look for
37 solutions to increase their resilience to a wide range of shocks and stresses (Mistry and Berardi 2016).
38 ILK can be deployed in the practice of climate governance especially at the local level where actions are
39 informed by the principles of decentralisation and autonomy (Chanza and de Wit 2016). ILK need not be
40 viewed as needing confirmation or disapproval by formal science, but rather it can complement scientific
41 knowledge (Klein et al. 2014).

42 The capacity to apply individual policy instruments and policy mixes is influenced by governance modes.
43 These modes include hierarchical governance that is centralised and imposes policy through top-down
44 measures, decentralised governance in which public policy is devolved to regional or local government,

1 public-private partnerships that aim for mutual benefits for the public and private sectors and self or
2 private governance that involves decisions beyond the realms of the public sector (IPBES 2018a). These
3 governance modes provide both constraints and opportunities for key actors that affect the effectiveness,
4 efficiency and equity of policy implementation.

5 **1.4.2 Gender agency as a critical factor in climate and land sustainability outcomes**

6 Environmental resource management is not gender neutral. Gender is an essential variable in shaping
7 ecological processes and change, building better prospects for livelihoods and sustainable development
8 (Resurrección 2013)(see Cross-Chapter Box 11: Gender, Chapter 7). Entrenched legal and social
9 structures and power relations constitute additional stressors that render women's experience of natural
10 resources disproportionately negative than men. Socio-economic drivers and entrenched gender
11 inequalities affect land-based management (Agarwal 2010). The intersections between climate change,
12 gender and climate adaptation takes place at multiple scales: household, national, international, and
13 adaptive capacities are shaped through power and knowledge.

14 Germaine to the gender inequities are the unequal access to land-based resources. Women play a
15 significant role in agriculture (Boserup 1989; Darity 1980) and rural economies globally (FAO 2011), but
16 are well below their share of labour in agriculture globally (FAO 2011). In 59% of 161 surveyed
17 countries, customary, traditional and religious practices hinder women land rights (OECD 2014).
18 Moreover, women typically shoulder disproportionate responsibility for unpaid domestic work including
19 care-giving activities (Beuchelt and Badstue 2013) and the provision of water and firewood (UNEP
20 2016). Exposure to violence restricts in large regions their mobility for capacity-building activities and
21 productive work outside the home (Day et al. 2005; UNEP 2016). Large-scale development projects can
22 erode rights, and lead to over-exploitation of natural resources. Hence, there are cases where reforms
23 related to land-based management, instead of enhancing food security, have tended to increase the
24 vulnerability of both women and men and reduce their ability to adapt to climate change (Pham et al.
25 2016). Access to, and control over, land and land-based resources is essential in taking concrete action to
26 land based mitigation, and inadequate access can affect women's rights and participation in land
27 governance and management of productive assets.

28 Timely information, such as from early warning systems, is critical in managing risks, disasters, and land
29 degradation, and in enabling land-based adaptation. Gender, household resources and social status, are all
30 determinants that influence the adoption of land-based strategies (Theriault et al. 2017). Climate change is
31 not a lone driver in the marginalisation of women, their ability to respond swiftly to its impacts will
32 depend on other socio-economic drivers that may help or hinder action towards adaptive governance.
33 Empowering women and removing gender-based inequities constitutes a mechanism for greater
34 participation in the adoption of sustainable practices of land management (Mello and Schmink 2017).
35 Improving women's access to land (Arora-Jonsson 2014) and other resources (water) and means of
36 economic livelihoods (such as credit and finance) are the prerequisites to enable women to participate in
37 governance and decision-making structures (Namubiru-Mwaura 2014). Still women are not a
38 homogenous group, and distinctions through elements of ethnicity, class, age and social status, require a
39 more nuanced approach and not a uniform treatment through vulnerability lenses only. An intersectional
40 approach that accounts for various social identifiers under different situation of power (Rao 2017) is
41 considered suitable to integrate gender into climate change research and helps to recognise overlapping
42 and interdependent systems of power (Djoudi et al. 2016; Kaijser and Kronsell 2014; Moosa and Tuana
43 2014; Thompson-Hall et al. 2016).

1 **1.4.3 Policy Instruments**

2 Policy instruments enable governance actors to respond to environmental and societal challenges through
3 policy action. Examples of the range of policy instruments available to public policy-makers is discussed
4 below based on four categories of instruments: legal and regulatory instruments, rights-based instruments
5 and customary norms, economic and financial instruments and social and cultural instruments.

6 **1.4.3.1 Legal & regulatory instruments**

7 Legal and regulatory instruments deal with all aspects of intervention by public policy organisations to
8 correct market failures, expand market reach, or intervene in socially relevant areas with inexistent
9 markets. Such instruments can include legislation to limit the impacts of intensive land management, for
10 example, protecting areas that are susceptible to nitrate pollution or soil erosion. Such instruments can
11 also set standards or threshold values, for example, mandated water quality limits, organic production
12 standards, or geographically defined regional food products. Legal and regulatory instruments may also
13 define liability rules, for example, where environmental standards are not met, as well as establishing
14 long-term agreements for land resource protection with land owners and land users.

15 **1.4.3.2 Economic and financial instruments**

16 Economic (such as taxes, subsidies) and financial (weather-index insurance) instruments deal with the
17 many ways in which public policy organisations can intervene in markets. A number of instruments are
18 available to support climate mitigation actions including public provision, environmental regulations,
19 creating property rights and markets, using markets (Sterner 2003). Market-based policies such as carbon
20 taxes, fuel taxes, cap and trade systems or green payments have been promoted (mostly in industrial
21 economies) to encourage markets and businesses to contribute to climate mitigation, but their
22 effectiveness to date has not always matched expectations (Grolleau et al. 2016) (see Section 7.4.4).
23 Market-based instruments in ecosystem services generate both positive (incentives for conservation), but
24 also negative environmental impacts, and also push food prices up or increase price instability (Gómez-
25 Baggethun and Muradian 2015; Farley and Voinov 2016). Footprint labels can be an effective means of
26 shifting consumer behaviour. However, private labels focusing on a single metric (e.g., carbon) may give
27 misleading signals if they target a portion of the life cycle (e.g., transport) (Appleton 2009) or ignore
28 other ecological indicators (water, nutrients, biodiversity)(van Noordwijk and Brussaard 2014).

29 Effective and durable, market-led responses for climate mitigation depend on business models that
30 internalise the cost of emissions into economic calculations. Such “business transformation” would itself
31 require integrated policies and strategies that aim to account for emissions in economic activities (Biagini
32 and Miller 2013; Weitzman 2014; Eidelwein et al. 2018). International initiatives such as REDD+ and
33 agricultural commodity roundtables (beef, soybeans, palm oil, sugar) are expanding the scope of private
34 sector participation in climate mitigation (Nepstad et al. 2013), but their impacts have not always been
35 effective (Denis et al. 2014). Payments for environmental services (PES) defined as “*voluntary*
36 *transactions between service users and service providers that are conditional on agreed rules of natural*
37 *resource management for generating offsite services*” (Wunder 2015) have not been widely adopted and
38 have not yet been demonstrated to deliver as effectively as originally hoped (Börner et al. 2017)(see
39 Sections 7.4, 7.5). PES in forestry were shown to be effective only when coupled with appropriate
40 regulatory measures (Alix-Garcia and Wolff 2014). Better designed and expanded PES schemes would
41 encourage integrated soil-water-nutrient management packages (Stavi et al. 2016), services for pollinator
42 protection (Nicole 2015), water use governance under scarcity and engage both public and private actors
43 (Loch et al. 2013). Effective PES also requires better economic metrics to account for human-directed

1 losses in terrestrial ecosystems and to food potential, and to address market failures or externalities
2 unaccounted for in market valuation of ecosystem services.

3 Resilient strategies for climate adaptation can rely on the construction of markets through social networks
4 as in the case of livestock systems (Denis et al. 2014) or when market signals encourage adaptation
5 through land markets or supply chain incentives for sustainable land management practices (Anderson et
6 al. 2018). Adequate policy (through regulations, investments in research and development or support to
7 social capabilities) can support private initiatives for effective solutions to restore degraded lands (Reed
8 and Stringer 2015), or mitigate against risk and to avoid shifting risks to the public (Biagini and Miller
9 2013). Governments, private business, and community groups could also partner to develop sustainable
10 production codes (Chartres and Noble 2015), and in co-managing land-based resources (Baker and
11 Chapin 2018), while private-public partnerships can be effective mechanisms in deploying infrastructure
12 to cope with climatic events (floods) and for climate-indexed insurance (Kunreuther 2015). Private
13 initiatives that depend on trade for climate adaptation and mitigation require reliable trading systems that
14 do not impede climate mitigation objectives (Elbehri et al 2015; Mathews 2017).

15 ***1.4.3.3 Rights-based instruments and customary norms***

16 Rights-based instruments and customary norms deal with the equitable and fair management of land
17 resources for all people (IPBES 2018a). These instruments emphasise the rights in particular of
18 indigenous peoples and local communities, including for example, recognition of the rights embedded in
19 the access to, and use of, common land. Common land includes situations without legal ownership (e.g.,
20 hunter-gathering communities in south America or Africa and bushmeat), where the legal ownership is
21 distinct from usage rights (Mediterranean transhumance grazing systems), or mixed ownership-common
22 grazing systems (e.g., Crofting in Scotland). A lack of formal (legal) ownership has often led to the loss
23 of access rights to land, where these rights were also not formally enshrined in law, which especially
24 affects indigenous communities, for example, deforestation in the Amazon basin. Overcoming the
25 constraints associated with common-pool resources (forestry, fisheries, water) are often of economic and
26 institutional nature (Hinkel et al. 2014) and require tackling the absence or poor functioning of institutions
27 and the structural constraints that they engender through access and control levers using policies and
28 markets and other mechanisms (Schut et al. 2016). Other examples of rights-based instruments include
29 the protection of heritage sites, sacred sites and peace parks (IPBES 2018a). Rights-based instruments and
30 customary norms are consistent with the aims of international and national human rights, and the critical
31 issue of liability in the climate change problem.

32 ***1.4.3.4 Social and cultural norms***

33 Social and cultural instruments are concerned with the communication of knowledge about conscious
34 consumption patterns and resource-effective ways of life through awareness raising, education and
35 communication of the quality and the provenance of land-based products. Examples of the latter include
36 consumption choices aided by ecolabelling (see 1.4.3.2) and certification. Cultural indicators (such as
37 social capital, cooperation, gender equity, women's knowledge, socio-ecological mobility) contribute to
38 the resilience of social-ecological systems (Sterling et al. 2017). Indigenous communities (such as the
39 Inuit and Tsleil Waututh Nation in Canada) that continue to maintain traditional foods exhibit greater
40 dietary quality and adequacy (Sheehy et al. 2015). Social and cultural instruments also include approaches
41 to self-regulation and voluntary agreements, especially with respect to environmental management and
42 land resource use. This is becoming especially important in the increasingly important domain of
43 corporate social responsibility (Halkos and Skouloudis 2016).

1 **1.5 The interdisciplinary nature of the SRCCL**

2 Assessing the land system in view of the multiple challenges that are covered by the SRCCL requires a
3 broad, inter-disciplinary perspective. Methods, core concepts and definitions are used differently in
4 different sectors, geographic regions, and across academic communities addressing land systems, and
5 these concepts and approaches to research are also undergoing a change in their interpretation through
6 time. These differences reflect varying perspectives, in nuances or emphasis, on land as component of the
7 climate and socio-economic systems. Because of its inter-disciplinary nature, the SRCCL can take
8 advantage of these varying perspectives and the diverse methods that accompany them. That way, the
9 report aims to support decision makers across sectors and world regions in the interpretation of its main
10 findings and support the implementation of solutions.

11

12 **Frequently Asked Questions**

13 **FAQ 1.1 What are the approaches to study the interactions between land and climate?**

14 Climate changes shapes the way land is able to support supply of food and water for humans. At the same
15 time the land surface interacts with the overlying atmosphere, thus human modifications of land use, land
16 cover and urbanisation affect global, regional and local climate. The complexity of the land-climate
17 interactions requires multiple study approaches embracing different spatial and temporal scales.
18 Observations of land atmospheric exchanges, such as of carbon, water, nutrients and energy can be
19 carried out at leaf level and soil with gas exchange systems, or at canopy scale by means of
20 micrometeorological techniques (i.e. eddy covariance). At regional scale, atmospheric measurements by
21 tall towers, aircraft and satellites can be combined with atmospheric transport models to obtain spatial
22 explicit maps of relevant greenhouse gases fluxes. At longer temporal scale (> 10 years) other approaches
23 are more effective such as tree ring chronologies, satellite records, population and vegetation dynamics
24 and isotopic studies. Models are important to bring information from measurement together and to extend
25 the knowledge in space and time, including the exploration of scenarios of future climate-land
26 interactions.

27

28 **FAQ 1.2 How region-specific are the impact of different land-based adaptation and mitigation 29 options?**

30 Land based adaptation and mitigation options are closely related to regional specific features for several
31 reasons. Climate change has a definite regional pattern with some regions already suffering from
32 enhanced climate extremes and others being impacted little, or even benefiting. From this point of view
33 increasing confidence in regional climate change scenarios is becoming a critical step forward towards the
34 implementation of adaptation and mitigation options. Biophysical and socio-economic impacts of climate
35 change depend on the exposures of natural ecosystems and economic sectors, which are again specific to
36 a region, reflecting regional sensitivities due to governance. The overall responses in terms of adaptation
37 or mitigation capacities to avoid and reduce vulnerabilities and enhance adaptive capacity, depend on
38 institutional arrangements, socio-economic conditions, and implementation of policies, many of them
39 having definite regional features. However global drivers, such as agricultural demand, food prices,
40 changing dietary habits associated with rapid social transformations (i.e. urban versus rural, meat versus

1 vegetarian) may interfere with regional specific policies for mitigation and adaptation options and require
2 the global level to be addressed.

3
4 **FAQ 1.3 What is the difference between desertification and land degradation? And where are they**
5 **happening?**

6 The difference between land degradation and desertification is geographic. Land degradation is a general
7 term used to describe a negative trend in land condition caused by direct or indirect human-induced
8 processes (including anthropogenic climate change). Degradation can be identified by the long-term
9 reduction or loss in biological productivity, ecological integrity or value to humans. Desertification is land
10 degradation when it occurs in arid, semi-arid, and dry sub-humid areas, which are also called drylands.
11 Contrary to some perceptions, desertification is not the same as the expansion of deserts. Desertification
12 is also not limited to irreversible forms of land degradation.

1 References

- 2 Abahussain, A. A., A. S. Abdu, W. K. Al-Zubari, N. A. El-Deen, and M. Abdul-Raheem, 2002:
3 Desertification in the Arab Region: Analysis of current status and trends. *J. Arid Environ.*, **51**, 521–
4 545, doi:10.1016/S0140-1963(02)90975-4.
5 <https://www.sciencedirect.com/science/article/pii/S0140196302909754> (Accessed April 25, 2018).
- 6 Abarca-Gómez, L., and Coauthors, 2017: Worldwide trends in body-mass index, underweight,
7 overweight, and obesity from 1975 to 2016: a pooled analysis of 2416 population-based
8 measurement studies in 128·9 million children, adolescents, and adults. *Lancet*, **390**, 2627–2642,
9 doi:10.1016/S0140-6736(17)32129-3.
10 <https://www.sciencedirect.com/science/article/pii/S0140673617321293> (Accessed April 14, 2019).
- 11 Abatzoglou, J. T., S. Z. Dobrowski, S. A. Parks, and K. C. Hegewisch, 2018: TerraClimate, a high-
12 resolution global dataset of monthly climate and climatic water balance from 1958–2015. *Sci. Data*,
13 **5**, 170191. <https://doi.org/10.1038/sdata.2017.191>.
- 14 Abberton, M., and Coauthors, 2016: Global agricultural intensification during climate change: a role for
15 genomics. *Plant Biotechnol. J.*, **14**, 1095–1098, doi:10.1111/pbi.12467.
16 <http://doi.wiley.com/10.1111/pbi.12467> (Accessed January 18, 2018).
- 17 Abreu, R. C. R., W. A. Hoffmann, H. L. Vasconcelos, N. A. Pilon, D. R. Rossatto, and G. Durigan, 2017:
18 The biodiversity cost of carbon sequestration in tropical savanna. *Sci. Adv.*, **3**, e1701284,
19 doi:10.1126/sciadv.1701284.
- 20 Abu Hammad, A., and A. Tumeizi, 2012: Land degradation: socioeconomic and environmental causes
21 and consequences in the eastern Mediterranean. *L. Degrad. Dev.*, **23**, 216–226,
22 doi:10.1002/ldr.1069. <http://doi.wiley.com/10.1002/ldr.1069> (Accessed April 22, 2018).
- 23 Achard, F., and Coauthors, 2014: Determination of tropical deforestation rates and related carbon losses
24 from 1990 to 2010. *Glob. Chang. Biol.*, **20**, 2540–2554, doi:10.1111/GCB.12605.
25 <https://onlinelibrary.wiley.com/doi/10.1111/gcb.12605/> (Accessed March 27, 2019).
- 26 Affognon, H., C. Mutungi, P. Sanginga, and C. Borgemeister, 2015: Unpacking Postharvest Losses in
27 Sub-Saharan Africa: A Meta-Analysis. *World Dev.*, **66**, 49–68,
28 doi:10.1016/J.WORLDDEV.2014.08.002.
29 <https://www.sciencedirect.com/science/article/pii/S0305750X14002307> (Accessed April 14, 2019).
- 30 Agarwal, B., 2010: *Gender and Green Governance*. Oxford University Press,
31 <http://www.oxfordscholarship.com/view/10.1093/acprof:oso/9780199569687.001.0001/acprof-9780199569687>
32 (Accessed April 3, 2019).
- 33 Agus, F., H. Husnain, and R. D. Yustika, 2015: Improve agricultural resilience to climate change through
34 soil management. *J. Penelit. dan Pengemb. Pertan.*, **34**, 147–158.
35 <http://ejurnal.litbang.pertanian.go.id/index.php/jppp/article/view/3094>.
- 36 Ahlstrom, A., G. Schurgers, A. Arneeth, and B. Smith, 2012: Robustness and uncertainty in terrestrial
37 ecosystem carbon response to CMIP5 climate change projections. *Environ. Res. Lett.*, **7**,
38 doi:10.1088/1748-9326/7/4/044008.
- 39 —, B. Smith, J. Lindstrom, M. Rummukainen, and C. B. Uvo, 2013: GCM characteristics explain the
40 majority of uncertainty in projected 21st century terrestrial ecosystem carbon balance.
41 *Biogeosciences*, **10**, 1517–1528, doi:10.5194/bg-10-1517-2013.
- 42 Ahrends, A., P. M. Hollingsworth, P. Beckschäfer, H. Chen, R. J. Zomer, L. Zhang, M. Wang, and J. Xu,
43 2017: China's fight to halt tree cover loss. *Proc. R. Soc. B Biol. Sci.*, **284**, 1–10,
44 doi:10.1098/rspb.2016.2559.

- 1 Aich, S., S. M. L. Ewe, B. Gu, T. W. Dreschel, S. Florida, W. Management, and W. P. Beach, 2014: An
2 evaluation of peat loss from an Everglades tree island , Florida , USA. *Mires Peat*, **14**, 1–15.
- 3 Albanito, F., T. Beringer, R. Corstanje, B. Poulter, A. Stephenson, J. Zawadzka, and P. Smith, 2016:
4 Carbon implications of converting cropland to bioenergy crops or forest for climate mitigation: a
5 global assessment. *GCB Bioenergy*, **8**, 81–95, doi:10.1111/gcbb.12242.
6 <http://doi.wiley.com/10.1111/gcbb.12242> (Accessed April 25, 2018).
- 7 Aleksandrowicz, L., R. Green, E. J. M. Joy, P. Smith, and A. Haines, 2016: The Impacts of Dietary
8 Change on Greenhouse Gas Emissions, Land Use, Water Use, and Health: A Systematic Review.
9 *PLoS One*, **11**, e0165797, doi:10.1371/journal.pone.0165797.
10 <http://dx.plos.org/10.1371/journal.pone.0165797> (Accessed January 18, 2018).
- 11 Alemu, M. M., 2016: Sustainable land management. *J. Environ. Prot. (Irvine,. Calif.)*, **7**, 502–506.
- 12 Alexander, P., D. Moran, M. D. A. Rounsevell, and P. Smith, 2013: Modelling the perennial energy crop
13 market: the role of spatial diffusion. *J. R. Soc. Interface*, **10**, in press, doi:10.1098/rsif.2013.0656.
14 <http://rsif.royalsocietypublishing.org/cgi/doi/10.1098/rsif.2013.0656>.
- 15 Alexander, P., M. D. A. Rounsevell, C. Dislich, J. R. Dodson, K. Engström, and D. Moran, 2015: Drivers
16 for global agricultural land use change: The nexus of diet, population, yield and bioenergy. *Glob.*
17 *Environ. Chang.*, doi:10.1016/j.gloenvcha.2015.08.011.
- 18 —, and Coauthors, 2016a: Assessing uncertainties in land cover projections. *Glob. Chang. Biol.*,
19 doi:10.1111/gcb.13447.
- 20 —, C. Brown, A. Arneeth, J. Finnigan, and M. D. A. Rounsevell, 2016b: Human appropriation of land
21 for food: The role of diet. *Glob. Environ. Chang. Policy Dimens.*, **41**, 88–98,
22 doi:10.1016/j.gloenvcha.2016.09.005.
- 23 —, —, —, —, D. Moran, and M. D. A. Rounsevell, 2017: Losses, inefficiencies and waste in
24 the global food syste. *Agric. Syst.*, **153**, 190–200, doi:10.1016/j.agry.2017.01.014.
25 <http://dx.doi.org/10.1016/j.agry.2017.01.014>.
- 26 —, S. Rabin, P. Anthoni, R. Henry, T. A. M. Pugh, M. D. A. Rounsevell, and A. Arneeth, 2018:
27 Adaptation of global land use and management intensity to changes in climate and atmospheric
28 carbon dioxide. *Glob. Chang. Biol.*, doi:10.1111/gcb.14110.
- 29 Alix-Garcia, J., and H. Wolff, 2014: Payment for Ecosystem Services from Forests. *Annu. Rev. Resour.*
30 *Econ.*, **6**, 361–380, doi:10.1146/annurev-resource-100913-012524.
31 <http://www.annualreviews.org/doi/10.1146/annurev-resource-100913-012524> (Accessed January 20,
32 2018).
- 33 Alkama, R., and A. Cescatti, 2016: Biophysical climate impacts of recent changes in global forest cover.
34 *Science (80-.)*, **351**, 600–604, doi:10.1126/science.aac8083.
- 35 Allan, T., M. Keulertz, and E. Woertz, 2015: The water-food-energy nexus: an introduction to nexus
36 concepts and some conceptual and operational problems (vol 31, pg 301, 2015). *Int. J. Water*
37 *Resour. Dev.*, **31**, 800, doi:10.1080/07900627.2015.1060725.
- 38 Allen, C. D., and Coauthors, 2010: A global overview of drought and heat-induced tree mortality reveals
39 emerging climate change risks for forests. *For. Ecol. Manage.*, **259**, 660–684,
40 doi:10.1016/j.foreco.2009.09.001.
- 41 Altieri, M. A., and C. I. Nicholls, 2017: The adaptation and mitigation potential of traditional agriculture
42 in a changing climate. *Clim. Change*, **140**, 33–45, doi:10.1007/s10584-013-0909-y.
- 43 Anav, A., and Coauthors, 2013: Evaluating the Land and Ocean Components of the Global Carbon Cycle
44 in the CMIP5 Earth System Models. *J. Clim.*, **26**, 6801–6843, doi:10.1175/jcli-d-12-00417.1.

- 1 Anderegg, W. R. L., J. M. Kane, and L. D. L. Anderegg, 2012: Consequences of widespread tree
2 mortality triggered by drought and temperature stress. *Nat. Clim. Chang.*, **3**, 30.
- 3 Anderegg, W. R. L., and Coauthors, 2015: Tree mortality from drought, insects, and their interactions in a
4 changing climate. *New Phytol.*, **208**, 674–683, doi:10.1111/nph.13477.
- 5 Anderson-Teixeira, K. J., 2018: Prioritizing biodiversity and carbon. *Nat. Clim. Chang.*, **8**, 667–668,
6 doi:10.1038/s41558-018-0242-6. <http://www.nature.com/articles/s41558-018-0242-6> (Accessed
7 March 21, 2019).
- 8 Anderson, K., and G. P. Peters, 2016: The trouble with negative emissions. *Science (80-.)*, **354**, 182–
9 183, doi:10.1126/science.aah4567.
- 10 Anderson, R. G., and Coauthors, 2011: Biophysical considerations in forestry for climate protection.
11 *Front. Ecol. Environ.*, **9**, 174–182, doi:10.1890/090179.
- 12 Anderson, S. E., and Coauthors, 2018: *The Critical Role of Markets in Climate Change Adaptation*.
13 National Bureau of Economic Research,.
- 14 Anseeuw, W., L. A. Wily, L. Cotula, and M. Taylor, 2011: *Land Rights and the Rush for Land*. 1-72 pp.
- 15 Ardö, J., T. Tagesson, S. Jamali, and A. Khatir, 2018: MODIS EVI-based net primary production in the
16 Sahel 2000–2014. *Int. J. Appl. Earth Obs. Geoinf.*, **65**, 35–45, doi:10.1016/j.jag.2017.10.002.
- 17 Arifanti, V. B., J. B. Kauffman, D. Hadriyanto, D. Murdiyarso, and R. Diana, 2018: Carbon dynamics and
18 land use carbon footprints in mangrove-converted aquaculture: The case of the Mahakam Delta,
19 Indonesia. *For. Ecol. Manage.*, **432**, 17–29, doi:10.1016/j.foreco.2018.08.047.
- 20 Arneth, A., C. Brown, and M. D. A. Rounsevell, 2014: Global models of human decision-making for
21 land-based mitigation and adaptation assessment. *Nat. Clim. Chang.*, **4**, 550–557,
22 doi:10.1038/nclimate2250. <http://www.nature.com/doi/finder/10.1038/nclimate2250>.
- 23 Arneth, A., and Coauthors, 2017: Historical carbon dioxide emissions caused by land-use changes are
24 possibly larger than assumed. *Nat. Geosci.*, **10**, 79, doi:10.1038/ngeo2882.
- 25 Aronson, M. F. J., and Coauthors, 2014: A global analysis of the impacts of urbanization on bird and
26 plant diversity reveals key anthropogenic drivers. *Proc. R. Soc. B Biol. Sci.*, **281**, 20133330–
27 20133330, doi:10.1098/rspb.2013.3330.
- 28 Arora-Jonsson, S., 2014: Forty years of gender research and environmental policy: Where do we stand?
29 *Womens. Stud. Int. Forum*, **47**, 295–308, doi:10.1016/J.WSIF.2014.02.009.
30 <https://www.sciencedirect.com/science/article/pii/S0277539514000326> (Accessed April 11, 2019).
- 31 Ashworth, K., O. Wild, and C. N. Hewitt, 2013: Impacts of biofuel cultivation on mortality and crop
32 yields. *Nat. Clim. Chang.*, **3**, 492–496, doi:10.1038/nclimate1788.
33 <http://dx.doi.org/10.1038/nclimate1788>.
- 34 Ayers, J. M., and S. Huq, 2009: The Value of Linking Mitigation and Adaptation: A Case Study of
35 Bangladesh. *Environ. Manage.*, **43**, 753–764, doi:10.1007/s00267-008-9223-2.
36 <http://link.springer.com/10.1007/s00267-008-9223-2> (Accessed April 3, 2019).
- 37 Azizi, A., A. Ghorbani, B. Malekmohammadi, and H. R. Jafari, 2017: Government management and
38 overexploitation of groundwater resources: absence of local community initiatives in Ardabil plain-
39 Iran. *J. Environ. Plan. Manag.*, **60**, 1785–1808, doi:10.1080/09640568.2016.1257975.
40 <https://doi.org/10.1080/09640568.2016.1257975>.
- 41 Bai, Z. G., D. L. Dent, L. Olsson, and M. E. Schaepman, 2008: Proxy global assessment of land
42 degradation. *Soil Use Manag.*, **24**, 223–234, doi:10.1111/j.1475-2743.2008.00169.x.
43 <http://doi.wiley.com/10.1111/j.1475-2743.2008.00169.x> (Accessed November 9, 2017).

- 1 Bais, A. L. S., C. Lauk, T. Kastner, and K. Erb, 2015: Global patterns and trends of wood harvest and use
2 between 1990 and 2010. *Ecol. Econ.*, **119**, 326–337, doi:10.1016/j.ecolecon.2015.09.011.
- 3 Bajželj, B., K. S. Richards, J. M. Allwood, P. Smith, J. S. Dennis, E. Curmi, and C. A. Gilligan, 2014:
4 Importance of food-demand management for climate mitigation. *Nat. Clim. Chang.*, **4**, 924,
5 doi:10.1038/nclimate2353. <http://dx.doi.org/10.1038/nclimate2353> (Accessed April 3, 2019).
- 6 Baker, S., and F. S. Chapin III, 2018: Going beyond" it depends:" the role of context in shaping
7 participation in natural resource management. *Ecol. Soc.*, **23**.
- 8 Baldos, U. L. C., and T. W. Hertel, 2015: The role of international trade in managing food security risks
9 from climate change. *Food Secur.*, **7**, 275–290, doi:10.1007/s12571-015-0435-z.
10 <http://link.springer.com/10.1007/s12571-015-0435-z> (Accessed April 22, 2018).
- 11 Baral, H., M. R. Guariguata, and R. J. Keenan, 2016: A proposed framework for assessing ecosystem
12 goods and services from planted forests. *Ecosyst. Serv.*, **22**, 260–268,
13 doi:10.1016/j.ecoser.2016.10.002. <http://dx.doi.org/10.1016/j.ecoser.2016.10.002>.
- 14 Bárcena, T. G., and Coauthors, 2014: Soil carbon stock change following afforestation in Northern
15 Europe: a meta-analysis. *Glob. Chang. Biol.*, **20**, 2393–2405, doi:10.1111/gcb.12576.
- 16 Barlow, J., and Coauthors, 2007: Quantifying the biodiversity value of tropical primary, secondary, and
17 plantation forests. *Proc. Natl. Acad. Sci. U. S. A.*, **104**, 18555–18560, doi:10.1073/pnas.0703333104.
18 <http://www.ncbi.nlm.nih.gov/pubmed/18003934> (Accessed August 17, 2018).
- 19 Barnett, J., and S. O'Neill, 2010: Maladaptation. *Glob. Environ. Chang.*, **2**, 211–213.
- 20 Bastin, J.-F., and Coauthors, 2017: The extent of forest in dryland biomes. *Science*, **356**, 635–638,
21 doi:10.1126/science.aam6527.
- 22 Bayer, A. D., M. Lindeskog, T. A. M. Pugh, P. M. Anthoni, R. Fuchs, and A. Arneeth, 2017: Uncertainties
23 in the land-use flux resulting from land-use change reconstructions and gross land transitions. *Earth*
24 *Syst. Dyn.*, **8**, 91–111, doi:10.5194/esd-8-91-2017.
- 25 Bazilian, M., and Coauthors, 2011: Considering the energy, water and food nexus: Towards an integrated
26 modelling approach. *Energy Policy*, **39**, 7896–7906, doi:10.1016/J.ENPOL.2011.09.039.
27 <https://www.sciencedirect.com/science/article/pii/S0301421511007282> (Accessed May 23, 2018).
- 28 Bebbber, D. P., and N. Butt, 2017: Tropical protected areas reduced deforestation carbon emissions by one
29 third from 2000-2012. *Sci. Rep.*, **7**, doi:1400510.1038/s41598-017-14467-w.
- 30 Beinroth, F. H., Eswaran, H., Reich, P. F. and Van Den Berg, E., 1994: *Stressed ecosystems and*
31 *sustainable agriculture*. Oxford & IBH Pub. C, New Dehli, 441 pp.
- 32 Bentsen, N. S., 2017: Carbon debt and payback time – Lost in the forest? *Renew. Sustain. Energy Rev.*,
33 **73**, 1211–1217, doi:10.1016/j.rser.2017.02.004.
34 <http://linkinghub.elsevier.com/retrieve/pii/S1364032117302034> (Accessed November 7, 2017).
- 35 Bernal, B., L. T. Murray, and T. R. H. Pearson, 2018: Global carbon dioxide removal rates from forest
36 landscape restoration activities. *Carbon Balance Manag.*, **13**, doi:10.1186/s13021-018-0110-8.
- 37 Berry, P. M., S. Brown, M. Chen, A. Kontogianni, O. Rowlands, G. Simpson, and M. Skourtos, 2015:
38 Cross-sectoral interactions of adaptation and mitigation measures. *Clim. Change*, **128**, 381–393,
39 doi:<http://dx.doi.org/10.1007/s10584-014-1214-0>.
40 <https://search.proquest.com/docview/1648241175?accountid=10673>.
- 41 Bertram, C., N. Johnson, G. Luderer, K. Riahi, M. Isaac, and J. Eom, 2015: Carbon lock-in through
42 capital stock inertia associated with weak near-term climate policies. *Technol. Forecast. Soc.*

- 1 *Change*, **90**, 62–72, doi:<https://doi.org/10.1016/j.techfore.2013.10.001>.
2 <http://www.sciencedirect.com/science/article/pii/S004016251300259X>.
- 3 Bestelmeyer, B. T., G. S. Okin, M. C. Duniway, S. R. Archer, N. F. Sayre, J. C. Williamson, and J. E.
4 Herrick, 2015: Desertification, land use, and the transformation of global drylands. *Front. Ecol.*
5 *Environ.*, **13**, 28–36, doi:[doi:10.1890/140162](https://doi.org/10.1890/140162).
6 <https://esajournals.onlinelibrary.wiley.com/doi/abs/10.1890/140162>.
- 7 Beuchelt, T. D., and L. Badstue, 2013: Gender, nutrition- and climate-smart food production:
8 Opportunities and trade-offs. *Food Secur.*, **5**, 709–721, doi:[10.1007/s12571-013-0290-8](https://doi.org/10.1007/s12571-013-0290-8).
9 <https://doi.org/10.1007/s12571-013-0290-8>.
- 10 Bhojvaid, P. P., M. P. Singh, S. R. Reddy, and J. Ashraf, 2016: Forest transition curve of India and related
11 policies, acts and other major factors. *Trop. Ecol.*, **57**, 133–141.
- 12 Biagini, B., and A. Miller, 2013: Engaging the private sector in adaptation to climate change in
13 developing countries: importance, status, and challenges. *Clim. Dev.*, **5**, 242–252,
14 doi:[10.1080/17565529.2013.821053](https://doi.org/10.1080/17565529.2013.821053). <https://doi.org/10.1080/17565529.2013.821053>.
- 15 Biggs, E. M., and Coauthors, 2015: Sustainable development and the water–energy–food nexus: A
16 perspective on livelihoods. *Environ. Sci. Policy*, **54**, 389–397, doi:[10.1016/J.ENVSCI.2015.08.002](https://doi.org/10.1016/J.ENVSCI.2015.08.002).
17 <https://www.sciencedirect.com/science/article/pii/S1462901115300563> (Accessed May 24, 2018).
- 18 Billen, G., L. Lassaletta, and J. Garnier, 2015: A vast range of opportunities for feeding the world in
19 2050: Trade-off between diet, N contamination and international trade. *Environ. Res. Lett.*, **10**,
20 doi:[10.1088/1748-9326/10/2/025001](https://doi.org/10.1088/1748-9326/10/2/025001).
- 21 Birdsey, R., and Y. Pan, 2015: Trends in management of the world’s forests and impacts on carbon
22 stocks. *For. Ecol. Manage.*, **355**, 83–90, doi:[10.1016/j.foreco.2015.04.031](https://doi.org/10.1016/j.foreco.2015.04.031).
23 <http://www.sciencedirect.com/science/article/pii/S0378112715002534> (Accessed September 27,
24 2016).
- 25 Bloom, A. A., J.-F. Exbrayat, I. R. van der Velde, L. Feng, and M. Williams, 2016: The decadal state of
26 the terrestrial carbon cycle: Global retrievals of terrestrial carbon allocation, pools, and residence
27 times. *Proc. Natl. Acad. Sci.*, **113**, 1285–1290, doi:[10.1073/pnas.1515160113](https://doi.org/10.1073/pnas.1515160113).
28 <http://www.pnas.org/content/113/5/1285> (Accessed February 22, 2016).
- 29 Bodin, Ö., 2017: Collaborative environmental governance: Achieving collective action in social-
30 ecological systems. *Science*, **357**, eaan1114, doi:[10.1126/science.aan1114](https://doi.org/10.1126/science.aan1114).
31 <http://www.ncbi.nlm.nih.gov/pubmed/28818915> (Accessed April 17, 2019).
- 32 Bodirsky, B. L., and C. Müller, 2014: Robust relationship between yields and nitrogen inputs indicates
33 three ways to reduce nitrogen pollution. *Environ. Res. Lett.*, doi:[10.1088/1748-9326/9/11/111005](https://doi.org/10.1088/1748-9326/9/11/111005).
- 34 Bonsch, M., and Coauthors, 2016: Trade-offs between land and water requirements for large-scale
35 bioenergy production. *GCB Bioenergy*, **8**, 11–24, doi:[10.1111/gcbb.12226](https://doi.org/10.1111/gcbb.12226).
36 [https://www.scopus.com/inward/record.uri?eid=2-s2.0-
37 84954285345%7B&%7DpartnerID=40%7B&%7Dmd5=37dc1171a883e87ad7283cdee3d2524c](https://www.scopus.com/inward/record.uri?eid=2-s2.0-84954285345%7B&%7DpartnerID=40%7B&%7Dmd5=37dc1171a883e87ad7283cdee3d2524c).
- 38 Börner, J., K. Baylis, E. Corbera, D. Ezzine-de-Blas, J. Honey-Rosés, U. M. Persson, and S. Wunder,
39 2017: The Effectiveness of Payments for Environmental Services. *World Dev.*, **96**, 359–374,
40 doi:[10.1016/J.WORLDDEV.2017.03.020](https://doi.org/10.1016/J.WORLDDEV.2017.03.020).
41 <https://www.sciencedirect.com/science/article/pii/S0305750X17300827> (Accessed November 2,
42 2018).
- 43 Borsato, E., P. Tarolli, and F. Marinello, 2018: Sustainable patterns of main agricultural products
44 combining different footprint parameters. *J. Clean. Prod.*, **179**, 357–367,

- 1 doi:10.1016/J.JCLEPRO.2018.01.044.
2 <https://www.sciencedirect.com/science/article/pii/S0959652618300556> (Accessed May 24, 2018).
- 3 Boserup, E., 1989: Population, the status of women, and rural development. *Popul. Dev. Rev.*, **15**, 45–60,
4 doi:10.2307/2807921. <https://www.jstor.org/stable/2807921>.
- 5 Boysen, L. R., W. Lucht, and D. Gerten, 2017: Trade-offs for food production, nature conservation and
6 climate limit the terrestrial carbon dioxide removal potential. *Glob. Chang. Biol.*, **23**, 4303–4317,
7 doi:10.1111/gcb.13745. <http://doi.wiley.com/10.1111/gcb.13745> (Accessed October 9, 2018).
- 8 Bradford, K. J., P. Dahal, J. Van Asbrouck, K. Kunusoth, P. Bello, J. Thompson, and F. Wu, 2018: The
9 dry chain: Reducing postharvest losses and improving food safety in humid climates. *Trends Food*
10 *Sci. Technol.*, **71**, 84–93, doi:10.1016/J.TIFS.2017.11.002.
11 <https://www.sciencedirect.com/science/article/pii/S092422441730482X> (Accessed April 14, 2019).
- 12 Bradshaw, C. J. A., N. S. Sodhi, K. S.-H. Peh, and B. W. Brook, 2007: Global evidence that deforestation
13 amplifies flood risk and severity in the developing world. *Glob. Chang. Biol.*, **13**, 2379–2395,
14 doi:10.1111/j.1365-2486.2007.01446.x.
- 15 Bren d'Amour, C., and Coauthors, 2016: Future urban land expansion and implications for global
16 croplands. *Proc. Natl. Acad. Sci.*, **114**, 201606036, doi:10.1073/pnas.1606036114.
17 <http://www.pnas.org/content/early/2016/12/20/1606036114> (Accessed January 9, 2017).
- 18 Brockerhoff, E. G., H. Jactel, J. A. Parrotta, and S. F. B. Ferraz, 2013: Role of eucalypt and other planted
19 forests in biodiversity conservation and the provision of biodiversity-related ecosystem services.
20 *For. Ecol. Manage.*, **301**, 43–50, doi:10.1016/j.foreco.2012.09.018.
21 <http://dx.doi.org/10.1016/j.foreco.2012.09.018>.
- 22 Brown, C., E. Brown, D. Murray-Rust, G. Cojocaru, C. Savin, and M. Rounsevell, 2014: Analysing
23 uncertainties in climate change impact assessment across sectors and scenarios. *Clim. Change*, **128**,
24 293–306, doi:10.1007/s10584-014-1133-0.
- 25 ———, P. Alexander, S. Holzhauser, and M. D. A. Rounsevell, 2017: Behavioral models of climate change
26 adaptation and mitigation in land-based sectors. *Wiley Interdiscip. Rev. Clim. Chang.*, **8**, e448,
27 doi:10.1002/wcc.448. <http://doi.wiley.com/10.1002/wcc.448>.
- 28 ———, ———, A. Arneth, I. Holman, and M. Rounsevell, 2019: Achievement of Paris climate goals unlikely
29 due to time lags in the land system. *Nat. Clim. Chang.*, **9**, 203–208, doi:10.1038/s41558-019-0400-
30 5. <https://doi.org/10.1038/s41558-019-0400-5>.
- 31 Brown, C. K., P. Alexander, and M. Rounsevell, 2018: Empirical evidence for the diffusion of knowledge
32 in land use change. *J. Land Use Sci.*,.
- 33 Brümmer, C., and Coauthors, 2017: Gas chromatography vs. quantum cascade laser-based N₂O flux
34 measurements using a novel chamber design. *Biogeosciences*, **14**, 1365–1381, doi:10.5194/bg-14-
35 1365-2017.
- 36 Bryan, B. A., and N. D. Crossman, 2013: Impact of multiple interacting financial incentives on land use
37 change and the supply of ecosystem services. *Ecosyst. Serv.*, **4**, 60–72,
38 doi:10.1016/j.ecoser.2013.03.004. <http://dx.doi.org/10.1016/j.ecoser.2013.03.004>.
- 39 Buisson, L., W. Thuiller, N. Casajus, S. Lek, and G. Grenouillet, 2009: Uncertainty in ensemble
40 forecasting of species distribution. *Glob. Chang. Biol.*, **16**, 1145–1157.
- 41 Bull, G. Q., M. Bazett, O. Schwab, S. Nilsson, A. White, and S. Maginnis, 2006: Industrial forest
42 plantation subsidies: Impacts and implications. *For. Policy Econ.*, **9**, 13–31,
43 doi:10.1016/j.forpol.2005.01.004.

- 1 Butler, M. P., P. M. Reed, K. Fisher-Vanden, K. Keller, and T. Wagener, 2014: Inaction and climate
2 stabilization uncertainties lead to severe economic risks. *Clim. Change*, **127**, 463–474,
3 doi:10.1007/s10584-014-1283-0. <http://link.springer.com/10.1007/s10584-014-1283-0> (Accessed
4 April 3, 2019).
- 5 Caffi, T., S. E. Legler, V. Rossi, and R. Bugiani, 2012: Evaluation of a Warning System for Early-Season
6 Control of Grapevine Powdery Mildew. *Plant Dis.*, **96**, 104–110, doi:10.1094/PDIS-06-11-0484.
7 <http://apsjournals.apsnet.org/doi/10.1094/PDIS-06-11-0484> (Accessed May 21, 2018).
- 8 Cafiero, C., S. Viviani, and M. Nord, 2018: Food security measurement in a global context: The food
9 insecurity experience scale. *Measurement*, **116**, 146–152,
10 doi:10.1016/J.MEASUREMENT.2017.10.065.
11 <https://www.sciencedirect.com/science/article/pii/S0263224117307005> (Accessed May 25, 2018).
- 12 Calvin, K., and B. Bond-Lamberty, 2018: Integrated human-earth system modeling - State of the science
13 and future directions. *Environ. Res. Lett.*, **13**, doi:10.1088/1748-9326/aac642.
- 14 Campbell, B. M., P. Thornton, R. Zougmore, P. van Asten, and L. Lipper, 2014: Sustainable
15 intensification: What is its role in climate smart agriculture? *Curr. Opin. Environ. Sustain.*, **8**, 39–
16 43, doi:<https://doi.org/10.1016/j.cosust.2014.07.002>.
17 <http://www.sciencedirect.com/science/article/pii/S1877343514000359>.
- 18 Canadell, J. G., and E. D. Schulze, 2014: Global potential of biospheric carbon management for climate
19 mitigation. *Nat. Commun.*, **5**, doi:528210.1038/ncomms6282.
- 20 Cao, S., J. Zhang, L. Chen, and T. Zhao, 2016: Ecosystem water imbalances created during ecological
21 restoration by afforestation in China, and lessons for other developing countries. *J. Environ.*
22 *Manage.*, **183**, 843–849, doi:10.1016/j.jenvman.2016.07.096.
23 <http://dx.doi.org/10.1016/j.jenvman.2016.07.096>.
- 24 Carlisle, K., and R. L. Gruby, 2017: Polycentric Systems of Governance: A Theoretical Model for the
25 Commons. *Policy Stud. J.*, doi:10.1111/psj.12212. <http://doi.wiley.com/10.1111/psj.12212>
26 (Accessed April 17, 2019).
- 27 Ceballos, G., P. R. Ehrlich, A. D. Barnosky, A. García, R. M. Pringle, and T. M. Palmer, 2015:
28 Accelerated modern human-induced species losses: Entering the sixth mass extinction. *Sci. Adv.*,
29 doi:10.1126/sciadv.1400253.
- 30 Cerretelli, S., and Coauthors, 2018: Spatial assessment of land degradation through key ecosystem
31 services: The role of globally available data. *Sci. Total Environ.*, **628–629**, 539–555,
32 doi:10.1016/J.SCITOTENV.2018.02.085.
33 <https://www.sciencedirect.com/science/article/pii/S0048969718304741> (Accessed April 22, 2018).
- 34 Challinor, A. J., J. Watson, D. B. Lobell, S. M. Howden, D. R. Smith, and N. Chhetri, 2014: A meta-
35 analysis of crop yield under climate change and adaptation. *Nat. Clim. Chang.*, **4**, 287–291,
36 doi:10.1038/nclimate2153.
- 37 Challinor, A. J., B. Parkes, and J. Ramirez-Villegas, 2015: Crop yield response to climate change varies
38 with cropping intensity. *Glob. Chang. Biol.*, **21**, 1679–1688, doi:10.1111/gcb.12808.
39 <http://doi.wiley.com/10.1111/gcb.12808> (Accessed January 18, 2018).
- 40 Challinor, A. J., W. N. Adger, T. G. Benton, D. Conway, M. Joshi, D. Frame, and A. J. Challinor, 2018:
41 Transmission of climate risks across sectors and borders Subject Areas : Author for correspondence :
- 42 Chanza, N., and A. de Wit, 2016: Enhancing climate governance through indigenous knowledge: Case in
43 sustainability science. *S. Afr. J. Sci.*, **Volume 112**, 1–7, doi:10.17159/sajs.2016/20140286.
44 <http://sajs.co.za/article/view/4058> (Accessed May 24, 2018).

- 1 Chappin, E. J. L., and T. van der Lei, 2014: Adaptation of interconnected infrastructures to climate
2 change: A socio-technical systems perspective. *Util. Policy*, **31**, 10–17,
3 doi:<https://doi.org/10.1016/j.jup.2014.07.003>.
4 <http://www.sciencedirect.com/science/article/pii/S0957178714000472>.
- 5 Chartres, C. J., and A. Noble, 2015: Sustainable intensification: overcoming land and water constraints on
6 food production. *Food Secur.*, **7**, 235–245, doi:10.1007/s12571-015-0425-1.
7 <https://doi.org/10.1007/s12571-015-0425-1>.
- 8 Chaturvedi, V., M. Hejazi, J. Edmonds, L. Clarke, P. Kyle, E. Davies, and M. Wise, 2015: Climate
9 mitigation policy implications for global irrigation water demand. *Mitig. Adapt. Strateg. Glob.*
10 *Chang.*, **20**, 389–407, doi:10.1007/s11027-013-9497-4. <https://doi.org/10.1007/s11027-013-9497-4>.
- 11 Chaudhary, A., and T. Kastner, 2016: Land use biodiversity impacts embodied in international food trade.
12 *Glob. Environ. Chang.*, **38**, 195–204, doi:10.1016/J.GLOENVCHA.2016.03.013.
13 <https://www.sciencedirect.com/science/article/pii/S0959378016300346> (Accessed May 24, 2018).
- 14 Chazdon, R. L., and Coauthors, 2016a: When is a forest a forest? Forest concepts and definitions in the
15 era of forest and landscape restoration. *Ambio*, doi:10.1007/s13280-016-0772-y.
- 16 Chazdon, R. L., and Coauthors, 2016b: Carbon sequestration potential of second-growth forest
17 regeneration in the Latin American tropics. *Sci. Adv.*, **2**, e1501639–e1501639,
18 doi:10.1126/sciadv.1501639.
- 19 Chegere, M. J., 2018: Post-harvest losses reduction by small-scale maize farmers: The role of handling
20 practices. *Food Policy*, **77**, 103–115, doi:10.1016/J.FOODPOL.2018.05.001.
21 <https://www.sciencedirect.com/science/article/abs/pii/S0306919217303706> (Accessed April 14,
22 2019).
- 23 Chen, B., and Coauthors, 2018: Global land-water nexus: Agricultural land and freshwater use embodied
24 in worldwide supply chains. *Sci. Total Environ.*, **613–614**, 931–943,
25 doi:10.1016/J.SCITOTENV.2017.09.138.
26 <https://www.sciencedirect.com/science/article/pii/S0048969717324877?via%3Dihub> (Accessed
27 April 3, 2019).
- 28 Chen, C., and Coauthors, 2019: China and India lead in greening of the world through land-use
29 management. *Nat. Sustain.*, **2**, 122–129, doi:10.1038/s41893-019-0220-7.
30 <http://www.nature.com/articles/s41893-019-0220-7>.
- 31 Chen, J., and Coauthors, 2014: Global land cover mapping at 30 m resolution: A POK-based operational
32 approach. *ISPRS J. Photogramm. Remote Sens.*, **103**, 7–27, doi:10.1016/j.isprsjprs.2014.09.002.
33 <http://dx.doi.org/10.1016/j.isprsjprs.2014.09.002>.
- 34 Chen, Y.-H., M. Babiker, S. Paltsev, and J. Reilly, 2016: *Costs of Climate Mitigation Policies*. MIT Joint
35 Program on the Science and Policy of Global Change,.
- 36 Cherlet, M., C. Hutchinson, J. Reynolds, J. Hill, S. Sommer, and G. von Maltitz, eds., 2018: *World atlas*
37 *of desertification: rethinking land degradation and sustainable land management*. Third edit.
38 Publication Office of the European Union, Luxembourg,.
- 39 Chipanshi, A., and Coauthors, 2015: Evaluation of the Integrated Canadian Crop Yield Forecaster
40 (ICCYF) model for in-season prediction of crop yield across the Canadian agricultural landscape.
41 *Agric. For. Meteorol.*, **206**, 137–150, doi:10.1016/J.AGRFORMET.2015.03.007.
42 <https://www.sciencedirect.com/science/article/pii/S0168192315000854?via%3Dihub> (Accessed
43 May 21, 2018).
- 44 Ciais, P., and Coauthors, 2013a: Carbon and Other Biogeochemical Cycles. *Climate Change 2013 - The*
45 *Physical Science Basis*, 465–570

- 1 http://www.ipcc.ch/report/ar5/wg1/docs/review/WG1AR5_SOD_Ch06_All_Final.pdf%5Cnhttp://ebooks.cambridge.org/ref/id/CBO9781107415324A023.
- 2
- 3 —, and Coauthors, 2013b: IPCC 5th Assessment Report: Working Group I: The physical science basis.
- 4 *IPCC Clim. Chang. Rep.*, doi:10.1017/CBO9781107415324.015.
- 5 Clark, D. A., S. Asao, R. Fisher, S. Reed, P. B. Reich, M. G. Ryan, T. E. Wood, and X. Yang, 2017:
- 6 Reviews and syntheses: Field data to benchmark the carbon cycle models for tropical forests.
- 7 *Biogeosciences*, **14**, 4663–4690, doi:10.5194/bg-14-4663-2017.
- 8 de Coninck, H., and Coauthors, 2018: Strengthening and implementing the global response. *Global*
- 9 *Warming of 1.5 °C: an IPCC special report on the impacts of global warming of 1.5C above pre-*
- 10 *industrial levels and related global greenhouse gas emission pathways, in the context of*
- 11 *strengthening the global response to the threat of climate change*, V. Masson-Delmotte et al., Eds.,
- 12 World Meteorological Organization, Geneva, Switzerland <http://www.ipcc.ch/report/sr15/>.
- 13 Constantin, C., C. Luminița, and A. J. Vasile, 2017: Land grabbing: A review of extent and possible
- 14 consequences in Romania. *Land use policy*, **62**, 143–150, doi:10.1016/j.landusepol.2017.01.001.
- 15 Costanza, R., R. de Groot, P. Sutton, S. van der Ploeg, S. J. Anderson, I. Kubiszewski, S. Farber, and R.
- 16 K. Turner, 2014: Changes in the global value of ecosystem services. *Glob. Environ. Chang.*, **26**,
- 17 152–158, doi:<https://doi.org/10.1016/j.gloenvcha.2014.04.002>.
- 18 <http://www.sciencedirect.com/science/article/pii/S0959378014000685>.
- 19 Cotula, L., and Coauthors, 2014: Testing Claims about Large Land Deals in Africa: Findings from a
- 20 Multi-Country Study. *J. Dev. Stud.*, **50**, 903–925, doi:10.1080/00220388.2014.901501.
- 21 Coyle, D. R., and Coauthors, 2017: Soil fauna responses to natural disturbances, invasive species, and
- 22 global climate change: Current state of the science and a call to action. *Soil Biol. Biochem.*, **110**,
- 23 116–133, doi:10.1016/J.SOILBIO.2017.03.008.
- 24 <https://www.sciencedirect.com/science/article/pii/S0038071717301530> (Accessed May 24, 2018).
- 25 Cremasch, G. D., 2016: *Sustainability metrics for agri-food supply chains*.
- 26 <http://library.wur.nl/WebQuery/wurpubs/fulltext/380247>.
- 27 Creutzig, F., and Coauthors, 2015: Bioenergy and climate change mitigation: an assessment. *Glob.*
- 28 *Chang. Biol. Bioenergy*, **7**, 916–944, doi:10.1111/gcbb.12205.
- 29 —, B. Fernandez, H. Haberl, R. Khosla, Y. Mulugetta, and K. C. Seto, 2016: Beyond Technology:
- 30 Demand-Side Solutions for Climate Change Mitigation. *Annu. Rev. Environ. Resour.*, **41**, 173–198,
- 31 doi:10.1146/annurev-environ-110615-085428. [http://www.annualreviews.org/doi/10.1146/annurev-](http://www.annualreviews.org/doi/10.1146/annurev-environ-110615-085428)
- 32 [environ-110615-085428](http://www.annualreviews.org/doi/10.1146/annurev-environ-110615-085428) (Accessed April 3, 2019).
- 33 —, and Coauthors, 2018: Towards demand-side solutions for mitigating climate change. *Nat. Clim.*
- 34 *Chang.*, **8**, 260–263, doi:10.1038/s41558-018-0121-1. [http://www.nature.com/articles/s41558-018-](http://www.nature.com/articles/s41558-018-0121-1)
- 35 [0121-1](http://www.nature.com/articles/s41558-018-0121-1) (Accessed April 11, 2019).
- 36 Crist, E., C. Mora, and R. Engelman, 2017: The interaction of human population, food production, and
- 37 biodiversity protection. *Science (80-.)*, **356**, 260–264, doi:10.1126/science.aal2011.
- 38 Crouzeilles, R., and Coauthors, 2016: A global meta-analysis on the ecological drivers of forest
- 39 restoration success. *Nat. Commun.*, **7**, 1–8, doi:1166610.1038/ncomms11666.
- 40 Cunningham, S. C., T. R. Cavagnaro, R. Mac Nally, K. I. Paul, P. J. Baker, J. Beringer, J. R. Thomson,
- 41 and R. M. Thompson, 2015a: Reforestation with native mixed-species plantings in a temperate
- 42 continental climate effectively sequesters and stabilizes carbon within decades. *Glob. Chang. Biol.*,
- 43 **21**, 1552–1566, doi:10.1111/gcb.12746.

- 1 Cunningham, S. C., R. Mac Nally, P. J. Baker, T. R. Cavagnaro, J. Beringer, J. R. Thomson, and R. M.
2 Thompson, 2015b: Balancing the environmental benefits of reforestation in agricultural regions.
3 *Perspect. Plant Ecol. Evol. Syst.*, **17**, 301–317, doi:10.1016/J.PPEES.2015.06.001.
4 <https://www.sciencedirect.com/science/article/pii/S1433831915000463> (Accessed September 19,
5 2018).
- 6 Curtis, P. G., C. M. Slay, N. L. Harris, A. Tyukavina, and M. C. Hansen, 2018: Classifying drivers of
7 global forest loss. *Science (80-.)*, **361**, 1108–1111, doi:10.1126/science.aau3445.
8 <http://www.sciencemag.org/lookup/doi/10.1126/science.aau3445>.
- 9 D’Odorico, P., A. Bhattachan, K. F. Davis, S. Ravi, and C. W. Runyan, 2013: Global desertification:
10 Drivers and feedbacks. *Adv. Water Resour.*, **51**, 326–344,
11 doi:<https://doi.org/10.1016/j.advwatres.2012.01.013>.
12 <http://www.sciencedirect.com/science/article/pii/S0309170812000231>.
- 13 D’Odorico, P., and Coauthors, 2018: The Global Food-Energy-Water Nexus. *Rev. Geophys.*, **56**, 456–
14 531, doi:10.1029/2017RG000591. <http://doi.wiley.com/10.1029/2017RG000591> (Accessed
15 November 2, 2018).
- 16 Daliakopoulos, I. N., I. K. Tsanis, A. Koutroulis, N. N. Kourgialas, A. E. Varouchakis, G. P. Karatzas,
17 and C. J. Ritsema, 2016: The threat of soil salinity: A European scale review. *Sci. Total Environ.*,
18 **573**, 727–739, doi:10.1016/J.SCITOTENV.2016.08.177.
19 <https://www.sciencedirect.com/science/article/pii/S0048969716318794> (Accessed May 24, 2018).
- 20 Dalin, C., and I. Rodríguez-Iturbe, 2016: Environmental impacts of food trade via resource use and
21 greenhouse gas emissions. *Environ. Res. Lett.*, **11**, 035012, doi:10.1088/1748-9326/11/3/035012.
22 [http://stacks.iop.org/1748-
23 9326/11/i=3/a=035012?key=crossref.aca35fdbf8a6626deefa7bd98d76a627](http://stacks.iop.org/1748-9326/11/i=3/a=035012?key=crossref.aca35fdbf8a6626deefa7bd98d76a627) (Accessed April 3,
24 2019).
- 25 Darity, W. A., 1980: The Boserup theory of agricultural growth: A model for anthropological economics.
26 *J. Dev. Econ.*, **7**, 137–157, doi:10.1016/0304-3878(80)90001-2.
27 <https://www.sciencedirect.com/science/article/pii/0304387880900012> (Accessed April 17, 2019).
- 28 Darrah, S. E., Y. Shennan-farpón, J. Loh, N. C. Davidson, C. M. Finlayson, R. C. Gardner, and M. J.
29 Walpole, 2019: Improvements to the Wetland Extent Trends (WET) index as a tool for monitoring
30 natural and human-made wetlands. *Ecol. Indic.*, **99**, 294–298, doi:10.1016/j.ecolind.2018.12.032.
31 <https://doi.org/10.1016/j.ecolind.2018.12.032>.
- 32 Davidson, N. C., 2014: How much wetland has the world lost? Long-term and recent trends in global
33 wetland area. *Mar. Freshw. Res.*, **65**, 934–941, doi:10.1071/MF14173.
34 <http://www.publish.csiro.au/?paper=MF14173> (Accessed October 7, 2018).
- 35 Davies-Barnard, T., P. J. Valdes, J. S. Singarayer, A. J. Wiltshire, and C. D. Jones, 2015: Quantifying the
36 relative importance of land cover change from climate and land-use in the representative
37 concentration pathway. *Global Biogeochem. Cycles*, 842–853,
38 doi:10.1002/2014GB004949.Received.
- 39 Davis, K. F., K. Yu, M. C. Rulli, L. Pichdara, and P. D’Odorico, 2015: Accelerated deforestation driven
40 by large-scale land acquisitions in Cambodia. *Nat. Geosci.*, **8**, 772–775, doi:10.1038/ngeo2540.
41 <http://www.nature.com/doi/10.1038/ngeo2540>.
- 42 Day, T., K. McKenna, and A. Bowlus, 2005: *The Economic Costs of Violence Against Women: An*
43 *Evaluation of the Literature. Expert brief compiled in preparation for the Secretary-General’s in-*
44 *depth study on all forms of violence against women.* NY: United Nations.Barzman, New York City,
45 1-66 pp.

- 1 Deininger, K., D. Byerlee, J. Lindsay, A. Norton, H. Selod, and M. Stickler, 2011: *Rising Global Interest*
2 *in Farmland, can it yield sustainable and equitable benefits?* 1st ed. The World Bank, Washington
3 D.C., 164 pp. http://siteresources.worldbank.org/INTARD/Resources/ESW_Sept7_final_final.pdf.
- 4 Dell'Angelo, J., P. D'Odorico, and M. C. Rulli, 2017a: Threats to sustainable development posed by land
5 and water grabbing. *Curr. Opin. Environ. Sustain.*, **26–27**, 120–128,
6 doi:10.1016/j.cosust.2017.07.007. <http://linkinghub.elsevier.com/retrieve/pii/S1877343517301756>.
- 7 ———, ———, ———, and P. Marchand, 2017b: The Tragedy of the Grabbed Commons: Coercion and
8 Dispossession in the Global Land Rush. *World Dev.*, **92**, 1–12,
9 doi:10.1016/J.WORLDDEV.2016.11.005.
10 <https://www.sciencedirect.com/science/article/pii/S0305750X15310445> (Accessed April 11, 2019).
- 11 Dendy, J., S. Cordell, C. P. Giardina, B. Hwang, E. Polloi, and K. Rengulbai, 2015: The role of remnant
12 forest patches for habitat restoration in degraded areas of Palau. *Restor. Ecol.*, **23**, 872–881,
13 doi:10.1111/rec.12268.
- 14 Deng, L., Z. Shangguan, and S. Sweeney, 2015: “Grain for Green” driven land use change and carbon
15 sequestration on the Loess Plateau, China. *Sci. Rep.*, **4**, 7039, doi:10.1038/srep07039.
16 <http://www.nature.com/articles/srep07039> (Accessed August 17, 2018).
- 17 Denis, G., L. Bruno, C. Christian, A. Véronique, and C. C. Gautier D, Locatelli B, 2014: Global changes,
18 livestock and vulnerability: the social construction of markets as an adaptive strategy. *Geogr. J.*,
19 **182**, 153–164, doi:10.1111/geoj.12115. [https://link.springer.com/article/10.1007/s12571-015-0425-](https://link.springer.com/article/10.1007/s12571-015-0425-1)
20 [1](https://link.springer.com/article/10.1007/s12571-015-0425-1).
- 21 Dietrich, J. P., and Coauthors, 2018: MAgPIE 4 - A modular open source framework for modeling global
22 land-systems. 1–26.
- 23 van Dijk, A. I. J. M., M. van Noordwijk, I. R. Calder, S. L. A. Bruijnzeel, J. Schellekens, and N. A.
24 Chappell, 2009: Forest – flood relation still tenuous – comment on ‘ Global evidence that
25 deforestation amplifies flood risk and severity in the developing world ’ by C . J . A . Bradshaw , N .
26 *Glob. Chang. Biol.*, **15**, 110–115, doi:10.1111/j.1365-2486.2008.01708.x.
- 27 Dinerstein, E., and Coauthors, 2015: Guiding Agricultural Expansion to Spare Tropical Forests. *Conserv.*
28 *Lett.*, **8**, 262–271, doi:10.1111/conl.12149.
- 29 Dixon, M. J. R., J. Loh, N. C. Davidson, C. Beltrame, R. Freeman, and M. Walpole, 2016: Tracking
30 global change in ecosystem area: The Wetland Extent Trends index. *Biol. Conserv.*,
31 doi:10.1016/j.biocon.2015.10.023.
- 32 Djoudi, H., B. Locatelli, C. Vaast, K. Asher, M. Brockhaus, and B. Basnett Sijapati, 2016: Beyond
33 dichotomies: Gender and intersecting inequalities in climate change studies. *Ambio*, **45**, 248–262,
34 doi:10.1007/s13280-016-0825-2. <http://link.springer.com/10.1007/s13280-016-0825-2> (Accessed
35 April 17, 2019).
- 36 Doelman, J. C., and Coauthors, 2018: Exploring SSP land-use dynamics using the IMAGE model:
37 Regional and gridded scenarios of land-use change and land-based climate change mitigation. *Glob.*
38 *Environ. Chang.*, **48**, 119–135, doi:<https://doi.org/10.1016/j.gloenvcha.2017.11.014>.
39 <http://www.sciencedirect.com/science/article/pii/S0959378016306392>.
- 40 Dohong, A., A. A. Aziz, and P. Dargusch, 2017: A review of the drivers of tropical peatland degradation
41 in South-East Asia. *Land use policy*, **69**, 349–360,
42 doi:<https://doi.org/10.1016/j.landusepol.2017.09.035>.
- 43 Dokken, D., 2014: *Part A: Global and Sectoral Aspects*. 1-1150 pp.
44 [papers2://publication/uuid/B8BF5043-C873-4AFD-97F9-A630782E590D](https://publication/uuid/B8BF5043-C873-4AFD-97F9-A630782E590D).

- 1 Donohue, R. J., M. L. Roderick, T. R. McVicar, and G. D. Farquhar, 2013: Impact of CO₂ fertilization on
2 maximum foliage cover across the globe's warm, arid environments. *Geophys. Res. Lett.*, **40**, 3031–
3 3035, doi:10.1002/grl.50563. <http://dx.doi.org/10.1002/grl.50563>.
- 4 Dooley, K., and S. Kartha, 2018: Land-based negative emissions: risks for climate mitigation and impacts
5 on sustainable development. *Int. Environ. Agreements Polit. Law Econ.*, **18**, 79–98,
6 doi:10.1007/s10784-017-9382-9. <https://doi.org/10.1007/s10784-017-9382-9>.
- 7 Doss, C., C. Kovarik, A. Peterman, A. Quisumbing, and M. van den Bold, 2015: Gender inequalities in
8 ownership and control of land in Africa: myth and reality. *Agric. Econ.*, **46**, 403–434,
9 doi:10.1111/agec.12171. <https://doi.org/10.1111/agec.12171>.
- 10 Dunford, R., P. A. Harrison, J. Jäger, M. D. A. Rounsevell, and R. Tinch, 2014: Exploring climate change
11 vulnerability across sectors and scenarios using indicators of impacts and coping capacity. *Clim.*
12 *Change*, **128**, 339–354, doi:10.1007/s10584-014-1162-8.
- 13 Duveiller, G., J. Hooker, and A. Cescatti, 2018: The mark of vegetation change on Earth's surface energy
14 balance. *Nat. Commun.*, **9**, 679, doi:10.1038/s41467-017-02810-8.
- 15 E. Appleton, A., 2009: *Private climate change standards and labelling schemes under the WTO*
16 *agreement on technical barriers to trade*. 131-152 pp.
- 17 Ebert, A. W., 2014: Potential of Underutilized Traditional Vegetables and Legume Crops to Contribute to
18 Food and Nutritional Security, Income and More Sustainable Production Systems. *Sustainability*, **6**,
19 319–335, doi:10.3390/su6010319. <http://www.mdpi.com/2071-1050/6/1/319/> (Accessed January 18,
20 2018).
- 21 Edenhofer, O., and Coauthors, 2014: Summary for Policymakers. *Climate Change 2014: Mitigation of*
22 *Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the*
23 *Intergovernmental Panel on Climate Change*, O. Edenhofer et al., Eds., Cambridge University
24 Press, Cambridge, United Kingdom and New York, NY, USA., p. 32.
- 25 Eraerts, M., I. Meeus, S. Van Den Berge, and G. Smagghe, 2017: Landscapes with high intensive fruit
26 cultivation reduce wild pollinator services to sweet cherry. *Agric. Ecosyst. Environ.*, **239**, 342–348,
27 doi:10.1016/J.AGEE.2017.01.031.
28 <https://www.sciencedirect.com/science/article/pii/S0167880917300415> (Accessed April 3, 2019).
- 29 Egginton, P., F. Beall, and J. Buttle, 2014: Reforestation - Climate change and water resource
30 implications. *For. Chron.*, **90**, 516–524, doi:10.5558/tfc2014-102.
- 31 Eidelwein, F., D. C. Collatto, L. H. Rodrigues, D. P. Lacerda, and F. S. Piran, 2018: Internalization of
32 environmental externalities: Development of a method for elaborating the statement of economic
33 and environmental results. *J. Clean. Prod.*, **170**, 1316–1327, doi:10.1016/J.JCLEPRO.2017.09.208.
- 34 Eitelberg, D. A., J. van Vliet, J. C. Doelman, E. Stehfest, and P. H. Verburg, 2016: Demand for
35 biodiversity protection and carbon storage as drivers of global land change scenarios. *Glob. Environ.*
36 *Chang.*, **40**, 101–111, doi:10.1016/j.gloenvcha.2016.06.014.
37 <http://linkinghub.elsevier.com/retrieve/pii/S0959378016300978>.
- 38 Elbehri, Aziz, Joshua Elliott, and T. W., 2015: *Climate change, food security and trade: an overview of*
39 *global assessments and policy insights*. 1-27 pp. http://centaur.reading.ac.uk/40644/1/FAO_1.pdf.
- 40 Elbehri, A., and Coauthors, 2017: *FAO-IPCC Expert Meeting on Climate Change, Land Use and Food*
41 *Security: Final Meeting Report*. Rome, 156 pp. [http://www.fao.org/documents/card/en/c/d5400b77-
42 1533-4c37-86a7-4945c320ea8d/](http://www.fao.org/documents/card/en/c/d5400b77-1533-4c37-86a7-4945c320ea8d/).
- 43 Ellis, E. C., and N. Ramankutty, 2008: Putting people in the map: anthropogenic biomes of the world.
44 *Front. Ecol. Environ.*, **6**, 439–447, doi:10.1890/070062. <https://doi.org/10.1890/070062>.

- 1 Ellis, E. C., K. K. Goldewijk, S. Siebert, D. Lightman, and N. Ramankutty, 2010: Anthropogenic
2 transformation of the biomes, 1700 to 2000. *Glob. Ecol. Biogeogr.*, doi:10.1111/j.1466-
3 8238.2010.00540.x.
- 4 Ellison, D., and Coauthors, 2017: Trees, forests and water: Cool insights for a hot world. *Glob. Environ.*
5 *Chang.*, **43**, 51–61, doi:https://doi.org/10.1016/j.gloenvcha.2017.01.002.
6 <http://www.sciencedirect.com/science/article/pii/S0959378017300134>.
- 7 Engstrom, K., S. Olin, M. D. A. Rounsevell, S. Brogaard, D. P. van Vuuren, P. Alexander, D. Murray-
8 Rust, and A. Arneth, 2016: Assessing uncertainties in global cropland futures using a conditional
9 probabilistic modelling framework. *Earth Syst. Dyn.*, **7**, 893–915, doi:10.5194/esd-7-893-2016.
- 10 Erb, K.-H., V. Gaube, F. Krausmann, C. Plutzer, A. Bondeau, and H. Haberl, 2007: A comprehensive
11 global 5 min resolution land-use data set for the year 2000 consistent with national census data. *J.*
12 *Land Use Sci.*, **2**, 191–224.
- 13 —, and Coauthors, 2016a: Land management: data availability and process understanding for global
14 change studies. *Glob. Chang. Biol.*, **23**, 512–533, doi:10.1111/gcb.13443.
15 <http://onlinelibrary.wiley.com/doi/10.1111/gcb.13443/abstract> (Accessed January 17, 2017).
- 16 —, C. Lauk, T. Kastner, A. Mayer, M. C. Theurl, and H. Haberl, 2016b: Exploring the biophysical
17 option space for feeding the world without deforestation. *Nat. Commun.*, **7**.
- 18 Erb, K.-H. H., H. Haberl, and C. Plutzer, 2012: Dependency of global primary bioenergy crop potentials
19 in 2050 on food systems, yields, biodiversity conservation and political stability. *Energy Policy*, **47**,
20 260–269, doi:10.1016/j.enpol.2012.04.066.
21 <https://www.sciencedirect.com/science/article/pii/S0301421512003710> (Accessed May 24, 2018).
- 22 —, and Coauthors, 2016c: Biomass turnover time in terrestrial ecosystems halved by land use. *Nat.*
23 *Geosci.*, **9**, 674–+, doi:10.1038/ngeo2782.
- 24 Erb, K.-H. K.-H. H. K.-H., and Coauthors, 2017: Unexpectedly large impact of forest management and
25 grazing on global vegetation biomass. *Nature*, **553**, 73–76, doi:10.1038/nature25138.
26 <https://www.nature.com/articles/nature25138> (Accessed December 21, 2017).
- 27 Evans, C. D., J. M. Williamson, F. Kacaribu, D. Irawan, Y. Suardiwerianto, M. Fikky, A. Laurén, and S.
28 E. Page, 2019: Rates and spatial variability of peat subsidence in Acacia plantation and forest
29 landscapes in Sumatra, Indonesia. *Geoderma*, **338**, 410–421, doi:10.1016/j.geoderma.2018.12.028.
- 30 Falco, S. Di, F. Adinolfi, M. Bozzola, and F. Capitanio, 2014: Crop Insurance as a Strategy for Adapting
31 to Climate Change. *J. Agric. Econ.*, **65**, 485–504, doi:doi:10.1111/1477-9552.12053.
32 <https://onlinelibrary.wiley.com/doi/abs/10.1111/1477-9552.12053>.
- 33 FAO's Animal Production and Health Division, 2018: Gridded Livestock of the World (GLW).
34 http://www.fao.org/Ag/againfo/resources/en/glw/GLW_prod-sys.html/. Accessed 17 October 2014.
- 35 FAO, 1963: *World Forest Inventory 1963*. Food and Agricultural Organization of the United Nations,
36 Rome, Italy, Italy, 113 pp.
- 37 —, 2011: *The State of Food and Agriculture: Women in agriculture - Closing the gender gap for*
38 *development*. Rome,.
- 39 —, 2015a: *Global Forest Resources Assessments 2015*. Food and Agriculture Organization of the
40 United Nations, Rome,.
- 41 —, 2015b: *Learning tool on Nationally Appropriate Mitigation Actions (NAMAs) in the agriculture,*
42 *forestry and other land use (AFOLU) sector*. 162 pp.

- 1 —, 2016: *State of the World's Forests 2016. Forests and agriculture: land-use challenges and*
2 *opportunities*. Rome, Italy, <http://www.fao.org/3/a-i5588e.pdf>.
- 3 —, 2017: *The future of food and agriculture: Trends and challenges*. Food and Agriculture
4 Organization of the United Nations, Rome,.
- 5 —, 2018a: *The State of the World's Forests 2018 - Forest pathways to sustainable development*. FAO,
6 Rome, 139 pp. <http://www.fao.org/publications/sofo/en/> (Accessed October 7, 2018).
- 7 —, 2018b: *The future of food and agriculture: Alternative pathways to 2050*. Food and Agricultural
8 Organization of the United Nations, Rome, 228 pp. <http://www.fao.org/publications/fofa/en/>
9 (Accessed October 8, 2018).
- 10 —, IFAD, UNICEF, WFP, and WHO, 2018: *The State of Food Security and Nutrition in the World*
11 *2018. Building climate resilience for food security and nutrition*. Rome,
12 <http://www.fao.org/3/I9553EN/i9553en.pdf>.
- 13 FAO and ITPS, 2015: *Status of the World's Soil Resources (SWSR)–Main Report*. Rome, Italy,
14 [https://xs.glgoo.net/scholar?hl=zh-](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=Status+of+the+World's+Soil+Resources+&btnG=)
15 [CN&as_sdt=0%2C5&q=Status+of+the+World's+Soil+Resources+&btnG=](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=Status+of+the+World's+Soil+Resources+&btnG=).
- 16 FAOSTAT, 2018: Statistical Databases. <http://faostat.fao.org> (Accessed March 22, 2016).
- 17 Farley, J., and A. Voinov, 2016: Economics, socio-ecological resilience and ecosystem services. *J.*
18 *Environ. Manage.*, **183**, 389–398, doi:10.1016/J.JENVMAN.2016.07.065.
19 <https://www.sciencedirect.com/science/article/pii/S0301479716305023> (Accessed April 17, 2019).
- 20 Fasullo, J. T., B. L. Otto-Bliesner, and S. Stevenson, 2018: ENSO's Changing Influence on Temperature,
21 Precipitation, and Wildfire In a Warming Climate. *Geophys. Res. Lett.*, **0**,
22 doi:doi:10.1029/2018GL079022.
23 <https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1029/2018GL079022>.
- 24 Favero, A., and E. Massetti, 2014: Trade of woody biomass for electricity generation under climate
25 mitigation policy. *Resour. Energy Econ.*, **36**, 166–190, doi:10.1016/J.RESENEECO.2013.11.005.
26 <https://www.sciencedirect.com/science/article/pii/S0928765513000808> (Accessed May 24, 2018).
- 27 —, and R. Mendelsohn, 2014: Using Markets for Woody Biomass Energy to Sequester Carbon in
28 Forests. *J. Assoc. Environ. Resour. Econ.*, **1**, 75–95, doi:10.1086/676033.
29 <http://www.jstor.org/stable/10.1086/676033>.
- 30 —, —, and B. Sohngen, 2017: Using forests for climate mitigation: sequester carbon or produce
31 woody biomass? *Clim. Change*, **144**, 195–206, doi:10.1007/s10584-017-2034-9.
32 <http://link.springer.com/10.1007/s10584-017-2034-9> (Accessed April 3, 2019).
- 33 Feng, X., and Coauthors, 2016: Revegetation in China's Loess Plateau is approaching sustainable water
34 resource limits. *Nat. Clim. Chang.*, **6**, 1019–1022, doi:10.1038/nclimate3092.
- 35 Fernandez-Martinez, M., and Coauthors, 2014: Nutrient availability as the key regulator of global forest
36 carbon balance. *Nat. Clim. Chang.*, **4**, 471–476, doi:10.1038/nclimate2177.
- 37 Ferreira, C. S. S., R. P. D. Walsh, and A. J. D. Ferreira, 2018: Degradation in urban areas. *Curr. Opin.*
38 *Environ. Sci. Heal.*, **5**, 19–25, doi:10.1016/j.coesh.2018.04.001.
39 <http://linkinghub.elsevier.com/retrieve/pii/S2468584417300570> (Accessed April 22, 2018).
- 40 Field, C. B., and Coauthors, 2014a: Summary for Policymakers. *Climate change 2014 : impacts,*
41 *adaptation, and vulnerability : Working Group II contribution to the fifth assessment report of the*
42 *Intergovernmental Panel on Climate Change*, Cambridge University Press, Cambridge, United
43 Kingdom and New York, NY, USA., 1–32.

- 1 Field, C. B., and Coauthors, 2014b: Technical Summary. *Climate Change 2014: Impacts, Adaptation, and*
2 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth*
3 *Assessment Report of the Intergovernmental Panel on Climate Change*, C.B. Field et al., Eds.,
4 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 35–94.
- 5 Filoso, S., M. O. Bezerra, K. C. B. Weiss, and M. A. Palmer, 2017: Impacts of forest restoration on water
6 yield: A systematic review. *PLoS One*, **12**, e0183210, doi:10.1371/journal.pone.0183210.
7 <http://dx.plos.org/10.1371/journal.pone.0183210> (Accessed September 19, 2018).
- 8 Fischer, J., and Coauthors, 2017: Reframing the Food-Biodiversity Challenge. *Trends Ecol. Evol.*, **32**,
9 335–345, doi:10.1016/j.tree.2017.02.009.
- 10 Fischer, M., and Coauthors, 2018: *IPBES: Summary for policymakers of the regional assessment report*
11 *on biodiversity and ecosystem services for Europe and Central Asia of the Intergovernmental*
12 *Science-Policy Platform on Biodiversity and Ecosystem Services*. Bonn, Germany, 48 pp.
- 13 Fish, R., A. Church, and M. Winter, 2016: Conceptualising cultural ecosystem services: A novel
14 framework for research and critical engagement. *Ecosyst. Serv.*, **21**, 208–217,
15 doi:<https://doi.org/10.1016/j.ecoser.2016.09.002>.
16 <http://www.sciencedirect.com/science/article/pii/S2212041616303138>.
- 17 Foley, J. A., and Coauthors, 2011: Solutions for a cultivated planet. *Nature*, **478**, 337–342,
18 doi:10.1038/nature10452.
- 19 Font Vivanco, D., R. Kemp, and E. van der Voet, 2016: How to deal with the rebound effect? A policy-
20 oriented approach. *Energy Policy*, **94**, 114–125, doi:10.1016/J.ENPOL.2016.03.054.
21 <https://www.sciencedirect.com/science/article/pii/S0301421516301586> (Accessed November 2,
22 2018).
- 23 Forsell, N., O. Turkovska, M. Gusti, M. Obersteiner, M. Elzen, and P. Havlík, 2016: *Assessing the*
24 *INDCs' land use, land use change, and forest emission projections*.
- 25 Franchini, M., and P. M. Mannucci, 2015: Impact on human health of climate changes. *Eur. J. Intern.*
26 *Med.*, **26**, 1–5, doi:10.1016/j.ejim.2014.12.008. <http://www.ncbi.nlm.nih.gov/pubmed/25582074>
27 (Accessed May 24, 2018).
- 28 Franco, A., and N. Giannini, 2005: Perspectives for the use of biomass as fuel in combined cycle power
29 plants. *Int. J. Therm. Sci.*, **44**, 163–177, doi:10.1016/J.IJTHERMALSCI.2004.07.005.
30 <https://www.sciencedirect.com/science/article/pii/S1290072904001875> (Accessed April 21, 2018).
- 31 Friedrich, T., R. Derpsch, and A. Kassam, 2012: Overview of the global spread of conservation
32 agriculture. *F. Actions Sci. Reports*, 1–7, doi:10.1201/9781315365800-4.
33 <http://factsreports.revues.org/1941>.
- 34 Friend, A. D., and Coauthors, 2014: Carbon residence time dominates uncertainty in terrestrial vegetation
35 responses to future climate and atmospheric CO₂. *Proc. Natl. Acad. Sci.*, **111**, 3280–3285,
36 doi:10.1073/pnas.1222477110. <http://www.pnas.org/lookup/doi/10.1073/pnas.1222477110>.
- 37 Friis, C., and J. Ø. Nielsen, 2017: Land-use change in a telecoupled world: The relevance and
38 applicability of the telecoupling framework in the case of banana plantation expansion in Laos.
39 *Ecol. Soc.*, doi:10.5751/ES-09480-220430.
- 40 —, —, I. Otero, H. Haberl, J. Niewöhner, and P. Hostert, 2016: From teleconnection to telecoupling:
41 taking stock of an emerging framework in land system science. *J. Land Use Sci.*,
42 doi:10.1080/1747423X.2015.1096423.

- 1 Fuchs, R., M. Herold, P. H. Verburg, J. G. P. W. Clevers, and J. Eberle, 2015: Gross changes in
2 reconstructions of historic land cover/use for Europe between 1900 and 2010. *Glob. Chang. Biol.*,
3 **21**, 299–313, doi:10.1111/gcb.12714.
- 4 ———, R. Prestele, and P. H. Verburg, 2017: A global assessment of gross and net land change dynamics
5 for current conditions and future scenarios. *Earth Syst. Dyn. Discuss.*, 1–29, doi:10.5194/esd-2017-
6 121.
- 7 Fuss, S., and Coauthors, 2018: Negative emissions—Part 2: Costs, potentials and side effects. *Environ.*
8 *Res. Lett.*, **13**, 063002, doi:10.1088/1748-9326/aabf9f. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/13/i=6/a=063002?key=crossref.280beee8a19ff00042252ae3ce163a06)
9 [9326/13/i=6/a=063002?key=crossref.280beee8a19ff00042252ae3ce163a06](http://stacks.iop.org/1748-9326/13/i=6/a=063002?key=crossref.280beee8a19ff00042252ae3ce163a06) (Accessed September
10 17, 2018).
- 11 Gaba, S., and Coauthors, 2015: Multiple cropping systems as drivers for providing multiple ecosystem
12 services: from concepts to design. *Agron. Sustain. Dev.*, **35**, 607–623, doi:10.1007/s13593-014-
13 0272-z. <http://link.springer.com/10.1007/s13593-014-0272-z> (Accessed November 2, 2018).
- 14 Gardiner, S. M., 2006: A Core Precautionary Principle*. *J. Polit. Philos.*, **14**, 33–60, doi:10.1111/j.1467-
15 9760.2006.00237.x. <http://doi.wiley.com/10.1111/j.1467-9760.2006.00237.x>.
- 16 Gerber, J. F., 2011: Conflicts over industrial tree plantations in the South: Who, how and why? *Glob.*
17 *Environ. Chang.*, **21**, 165–176, doi:10.1016/j.gloenvcha.2010.09.005.
- 18 Gibbs, H. K., and J. M. Salmon, 2015: Mapping the world’s degraded lands. *Appl. Geogr.*, **57**, 12–21,
19 doi:10.1016/j.apgeog.2014.11.024.
20 <https://www.sciencedirect.com/science/article/pii/S0143622814002793> (Accessed April 25, 2018).
- 21 Gilbert-Norton, L., R. Wilson, J. R. Stevens, and K. H. Beard, 2010: A Meta-Analytic Review of Corridor
22 Effectiveness. *Conserv. Biol.*, **24**, 660–668, doi:10.1111/j.1523-1739.2010.01450.x.
- 23 Gillett, N. P., and Coauthors, 2016: The Detection and Attribution Model Intercomparison Project
24 (DAMIP v1.0) contribution to CMIP6. *Geosci. Model Dev.*, **9**, 3685–3697, doi:10.5194/gmd-9-
25 3685-2016.
- 26 Gillingham, K., and J. H. Stock, 2018: The Cost of Reducing Greenhouse Gas Emissions. *J. Econ.*
27 *Perspect.*, **32**, 53–72.
- 28 Gilmont, M., 2015: Water resource decoupling in the MENA through food trade as a mechanism for
29 circumventing national water scarcity. *Food Secur.*, **7**, 1113–1131, doi:10.1007/s12571-015-0513-2.
30 <https://doi.org/10.1007/s12571-015-0513-2>.
- 31 Gleckler P. J., Doutriaux C., Durack P. J., Taylor K. E., Zhang Y., Williams D. N., Mason E., and S. J.,
32 2016: A more powerful reality test for climate models. *Eos (Washington. DC)*, **97**,
33 doi:10.1029/2016EO051663.
- 34 Goldewijk, K. K., A. Beusen, J. Doelman, and E. Stehfest, 2017: Anthropogenic land use estimates for
35 the Holocene - HYDE 3.2. *Earth Syst. Sci. Data*, **9**, 927–953, doi:10.5194/essd-9-927-2017.
- 36 Goldstein, A., W. R. Turner, J. Gladstone, and D. G. Hole, 2019: The private sector’s climate change risk
37 and adaptation blind spots. *Nat. Clim. Chang.*, **9**, 18–25, doi:10.1038/s41558-018-0340-5.
38 <http://www.nature.com/articles/s41558-018-0340-5> (Accessed April 11, 2019).
- 39 Golub, A. A., B. B. Henderson, T. W. Hertel, P. J. Gerber, S. K. Rose, and B. Sohngen, 2013: Global
40 climate policy impacts on livestock, land use, livelihoods, and food security. *Proc. Natl. Acad. Sci.*
41 *U. S. A.*, **110**, 20894–20899, doi:10.1073/pnas.1108772109.
42 <http://www.ncbi.nlm.nih.gov/pubmed/23019587> (Accessed April 3, 2019).
- 43 Gómez-Baggethun, E., and R. Muradian, 2015: In markets we trust? Setting the boundaries of Market-
44 Based Instruments in ecosystem services governance. *Ecol. Econ.*, **117**, 217–224,

- 1 doi:10.1016/J.ECOLECON.2015.03.016.
2 <https://www.sciencedirect.com/science/article/pii/S0921800915001019> (Accessed April 17, 2019).
- 3 Gonzalez, P., R. P. Neilson, J. M. Lenihan, and R. J. Drapek, 2010: Global patterns in the vulnerability of
4 ecosystems to vegetation shifts due to climate change. *Glob. Ecol. Biogeogr.*, **19**, 755–768,
5 doi:10.1111/j.1466-8238.2010.00558.x.
- 6 Gossner, M. M., and Coauthors, 2016: Land-use intensification causes multitrophic homogenization of
7 grassland communities. *Nature*, doi:10.1038/nature20575.
- 8 Graham, C. T., M. W. Wilson, T. Gittings, T. C. Kelly, S. Irwin, J. L. Quinn, and J. O’Halloran, 2017:
9 Implications of afforestation for bird communities: the importance of preceding land-use type.
10 *Biodivers. Conserv.*, **26**, 3051–3071, doi:10.1007/s10531-015-0987-4.
- 11 Grassi, G., J. House, F. Dentener, S. Federici, M. den Elzen, and J. Penman, 2017: The key role of forests
12 in meeting climate targets requires science for credible mitigation. *Nat. Clim. Chang.*, **7**, 220–+,
13 doi:10.1038/nclimate3227. [https://xs.glgoo.net/scholar?hl=zh-](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=The+key+role+of+forests+in+meeting+climate+targets+requires+science+for+credible+mitigation&btnG=)
14 [CN&as_sdt=0%2C5&q=The+key+role+of+forests+in+meeting+climate+targets+requires+science+](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=The+key+role+of+forests+in+meeting+climate+targets+requires+science+for+credible+mitigation&btnG=)
15 [for+credible+mitigation&btnG=.](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=The+key+role+of+forests+in+meeting+climate+targets+requires+science+for+credible+mitigation&btnG=)
- 16 Griffith, D. M., and Coauthors, 2017: Comment on “The extent of forest in dryland biomes.” *Science* (80-
17 .), **358**, eaao1309, doi:10.1126/science.aao1309. <http://www.ncbi.nlm.nih.gov/pubmed/29146777>
18 (Accessed September 19, 2018).
- 19 Griscom, B. W., and Coauthors, 2017: Natural climate solutions. *Proc. Natl. Acad. Sci. U. S. A.*, **114**,
20 11645–11650, doi:10.1073/pnas.1710465114.
- 21 Grolleau, G., L. Ibanez, N. Mzoughi, and M. Teisl, 2016: Helping eco-labels to fulfil their promises.
22 *Clim. Policy*, **16**, 792–802, doi:10.1080/14693062.2015.1033675.
23 <http://www.tandfonline.com/doi/full/10.1080/14693062.2015.1033675> (Accessed November 2,
24 2018).
- 25 de Groot, W. J., B. M. Wotton, and M. D. Flannigan, 2015: Chapter 11 - Wildland Fire Danger Rating
26 and Early Warning Systems. *Wildfire Hazards, Risks and Disasters*, J.F. Shroder and D. Paton, Eds.,
27 Elsevier, Oxford, 207–228
28 <http://www.sciencedirect.com/science/article/pii/B9780124104341000117>.
- 29 Güneralp, B., K. C. Seto, B. Gueneralp, and K. C. Seto, 2013: Futures of global urban expansion:
30 Uncertainties and implications for biodiversity conservation. *Environ. Res. Lett.*, **8**,
31 doi:10.1088/1748-9326/8/1/014025.
- 32 Gurwick, N. P., L. A. Moore, C. Kelly, and P. Elias, 2013: A Systematic Review of Biochar Research,
33 with a Focus on Its Stability in situ and Its Promise as a Climate Mitigation Strategy. *PLoS One*, **8**,
34 doi:10.1371/journal.pone.0075932.
- 35 Gustavsson, J., C. Cederberg, U. Sonesson, R. van Otterdijk, and A. Meybeck, 2011: Global Food Losses
36 and Food Waste—Extent, Causes and Prevention.
- 37 Haasnoot, M., 2013: Dynamic adaptive policy pathways: A method for crafting robust decisions for a
38 deeply uncertain world. *Glob. Environ. Chang.*, **23**, 485–498, doi:10.1016/j.gloenvcha.2012.12.006.
39 <https://www.sciencedirect.com/science/article/pii/S095937801200146X>.
- 40 Haberl, H., 2015: *Competition for land: A sociometabolic perspective*. Elsevier,
41 <http://www.sciencedirect.com/science/article/pii/S0921800914003127> (Accessed November 20,
42 2015).
- 43 ———, K.-H. K.-H. K.-H. Erb, and F. Krausmann, 2014: Human Appropriation of Net Primary Production:
44 Patterns, Trends, and Planetary Boundaries. *Annu. Rev. Environ. Resour.*, **39**, 363–391,

- 1 doi:10.1146/annurev-environ-121912-094620. <http://www.annualreviews.org/doi/10.1146/annurev-environ-121912-094620>.
- 2
- 3 Haddad, N. M., and Coauthors, 2015: Habitat fragmentation and its lasting impact on Earth’s ecosystems.
4 *Sci. Adv.*, **1**, doi:10.1126/sciadv.1500052.
5 <http://advances.sciencemag.org/content/advances/1/2/e1500052.full.pdf>.
- 6 Halkos, G., and A. Skouloudis, 2016: Cultural dimensions and corporate social responsibility: A cross-
7 country analysis. <https://mpra.ub.uni-muenchen.de/69222/>.
- 8 Hallegatte, S., and J. Rentschler, 2015: Risk Management for Development-Assessing Obstacles and
9 Prioritizing Action. *Risk Anal.*, **35**, 193–210, doi:10.1111/risa.12269.
- 10 —, and K. J. Mach, 2016: Make climate-change assessments more relevant. *Nature*, **534**, 613–615,
11 doi:10.1038/534613a.
- 12 Hallström, E., A. Carlsson-Kanyama, and P. Börjesson, 2015: Environmental impact of dietary change: a
13 systematic review. *J. Clean. Prod.*, **91**, 1–11, doi:10.1016/J.JCLEPRO.2014.12.008.
14 <https://www.sciencedirect.com/science/article/pii/S0959652614012931> (Accessed October 22,
15 2018).
- 16 Hansen, M. C., and Coauthors, 2013: High-Resolution Global Maps of 21st-Century Forest Cover
17 Change. *Science (80-.)*, **342**, 850–853, doi:10.1126/science.1244693.
- 18 Harper, A. B., and Coauthors, 2018: Land-use emissions play a critical role in land-based mitigation for
19 Paris climate targets. *Nat. Commun.*, **9**, doi:10.1038/s41467-018-05340-z.
- 20 Harrison, P. A., R. Dunford, C. Savin, M. D. A. Rounsevell, I. P. Holman, A. S. Kebede, and B. Stuch,
21 2014: Cross-sectoral impacts of climate change and socio-economic change for multiple, European
22 land- and water-based sectors. *Clim. Change*, **128**, 279–292, doi:10.1007/s10584-014-1239-4.
- 23 Harrison, P. A., R. W. Dunford, I. P. Holman, and M. D. A. Rounsevell, 2016: Climate change impact
24 modelling needs to include cross-sectoral interactions. *Nat. Clim. Chang.*, **6**, 885–890,
25 doi:10.1038/nclimate3039.
- 26 Harrison, S. P., and Coauthors, 2013: Volatile isoprenoid emissions from plastid to planet. *New Phytol.*,
27 **197**, 49–57, doi:10.1111/nph.12021.
- 28 Harvey, M., and S. Pilgrim, 2011: The new competition for land: Food, energy, and climate change. *Food*
29 *Policy*, **36**, S40–S51, doi:10.1016/J.FOODPOL.2010.11.009.
30 <https://www.sciencedirect.com/science/article/pii/S0306919210001235> (Accessed May 26, 2018).
- 31 Hasegawa, T., S. Fujimori, Y. Shin, A. Tanaka, K. Takahashi, and T. Masui, 2015: Consequence of
32 Climate Mitigation on the Risk of Hunger. *Environ. Sci. Technol.*, **49**, 7245–7253,
33 doi:10.1021/es5051748.
- 34 —, and Coauthors, 2018: Risk of increased food insecurity under stringent global climate change
35 mitigation policy. *Nat. Clim. Chang.*, **8**, 699–703, doi:10.1038/s41558-018-0230-x.
36 <http://dx.doi.org/10.1038/s41558-018-0230-x>.
- 37 Hayley, L., C. Declan, B. Michael, and R. Judith, 2015: Tracing the Water–Energy–Food Nexus:
38 Description, Theory and Practice. *Geogr. Compass*, **9**, 445–460, doi:10.1111/gec3.12222.
39 <https://doi.org/10.1111/gec3.12222>.
- 40 He, T., S. Liang, D. Wang, Y. Cao, F. Gao, Y. Yu, and M. Feng, 2018: Evaluating land surface albedo
41 estimation from Landsat MSS, TM, ETM+, and OLI data based on the unified direct estimation
42 approach. *Remote Sens. Environ.*, **204**, 181–196, doi:10.1016/j.rse.2017.10.031.

- 1 Heck, V., D. Gerten, W. Lucht, and A. Popp, 2018: Biomass-based negative emissions difficult to
2 reconcile with planetary boundaries. *Nat. Clim. Chang.*, **8**, 151–155, doi:10.1038/s41558-017-0064-
3 y. <https://doi.org/10.1038/s41558-017-0064-y>.
- 4 Heilmayr, R., C. Echeverría, R. Fuentes, and E. F. Lambin, 2016: A plantation-dominated forest
5 transition in Chile. *Appl. Geogr.*, **75**, 71–82, doi:10.1016/j.apgeog.2016.07.014.
6 <http://dx.doi.org/10.1016/j.apgeog.2016.07.014>.
- 7 Hennessey, R., J. Pittman, A. Morand, and A. Douglas, 2017: Co-benefits of integrating climate change
8 adaptation and mitigation in the Canadian energy sector. *Energy Policy*, **111**, 214–221,
9 doi:10.1016/J.ENPOL.2017.09.025.
10 <https://www.sciencedirect.com/science/article/pii/S0301421517305852> (Accessed April 11, 2019).
- 11 Henry, R. C., and Coauthors, 2018: Food supply and bioenergy production within the global cropland
12 planetary boundary. *PLoS One*, **13**, e0194695–e0194695, doi:10.1371/journal.pone.0194695.
13 <http://dx.plos.org/10.1371/journal.pone.0194695> (Accessed May 24, 2018).
- 14 Hernández-Morcillo, M., T. Plieninger, and C. Bieling, 2013: An empirical review of cultural ecosystem
15 service indicators. *Ecol. Indic.*, **29**, 434–444, doi:<https://doi.org/10.1016/j.ecolind.2013.01.013>.
16 <http://www.sciencedirect.com/science/article/pii/S1470160X13000320>.
- 17 Hersperger, A. M., M.-P. Gennaio, P. H. Verburg, xfc, and M. rgi, 2010: Linking Land Change with
18 Driving Forces and ActorsFour Conceptual Models. *Ecol. Soc.*, **15**.
19 <http://www.jstor.org/stable/26268195>.
- 20 Hinkel, J., P. W. G. Bots, and M. Schlöter, 2014: Enhancing the Ostrom social-ecological system
21 framework through formalization. *Ecol. Soc.*, **19**. <http://www.jstor.org/stable/26269623>.
- 22 HLPE, 2017: *Nutrition and Food Systems. A report by the High Level Panel of Experts on Food Security*
23 *and Nutrition of the Committee on World Food Security*. Rome.,.
- 24 Hobbs, P. R., K. Sayre, and R. Gupta, 2008: The role of conservation agriculture in sustainable
25 agriculture. *Philos. Trans. R. Soc. B Biol. Sci.*, **363**, 543–555, doi:10.1098/rstb.2007.2169.
26 <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2007.2169>.
- 27 Hoegh-Guldberg, O., and Coauthors, 2018: Impacts of 1.5°C Global Warming on Natural and Human
28 Systems. *Global warming of 1.5°C. An IPCC Special Report on the impacts of global warming of*
29 *1.5°C above pre-industrial levels and related global greenhouse gas emission pathways, in the*
30 *context of strengthening the global response to the threat of climate change*, V. Masson-Delmotte et
31 al., Eds., World Meteorological Organization, Geneva, Switzerland.
- 32 Hoekstra, A. Y., and T. O. Wiedmann, 2014: Humanity’s unsustainable environmental footprint. *Science*
33 *(80-)*, **344**, 1114–1117, doi:10.1126/science.1248365.
- 34 Hoesly, R. M., and Coauthors, 2018: Historical (1750–2014) anthropogenic emissions of reactive gases
35 and aerosols from the Community Emissions Data System (CEDS). *Geosci. Model Dev.*, **11**, 369–
36 408, doi:10.5194/gmd-11-369-2018. <https://www.geosci-model-dev.net/11/369/2018/>.
- 37 Hof, C., A. Voskamp, M. F. Biber, K. Böhning-Gaese, E. K. Engelhardt, A. Niamir, S. G. Willis, and T.
38 Hickler, 2018: Bioenergy cropland expansion may offset positive effects of climate change
39 mitigation for global vertebrate diversity. *Proc. Natl. Acad. Sci.*, **115**, 13294–13299,
40 doi:10.1073/pnas.1807745115. <https://www.pnas.org/content/pnas/115/52/13294.full.pdf>.
- 41 Hoff, H., 2011: *Bonn2011 Conference The Water, Energy and Food Security Nexus Solutions for the*
42 *Green Economy*. Stockholm, 1-52 pp.
- 43 Houghton, R. A., 2013: The emissions of carbon from deforestation and degradation in the tropics: Past
44 trends and future potential. *Carbon Manag.*, **4**, 539–546, doi:10.4155/cmt.13.41.

- 1 —, and A. A. Nassikas, 2017: Global and regional fluxes of carbon from land use and land cover
2 change 1850–2015. *Global Biogeochem. Cycles*, **31**, 456–472, doi:10.1002/2016GB005546.
- 3 —, B. Byers, and A. A. Nassikas, 2015: A role for tropical forests in stabilizing atmospheric CO₂. *Nat.*
4 *Clim. Chang.*, **5**, 1022–1023, doi:10.1038/nclimate2869. <http://dx.doi.org/10.1038/nclimate2869>.
- 5 Howells, M., and Coauthors, 2013: Integrated analysis of climate change, land-use, energy and water
6 strategies. *Nat. Clim. Chang.*, **3**, 621. <http://dx.doi.org/10.1038/nclimate1789>.
- 7 Hua, F., and Coauthors, 2016: Opportunities for biodiversity gains under the world’s largest reforestation
8 programme. *Nat. Commun.*, **7**, 1–11, doi:10.1038/ncomms12717.
9 <http://dx.doi.org/10.1038/ncomms12717>.
- 10 —, and Coauthors, 2018: Tree plantations displacing native forests: The nature and drivers of apparent
11 forest recovery on former croplands in Southwestern China from 2000 to 2015. *Biol. Conserv.*, **222**,
12 113–124, doi:10.1016/j.biocon.2018.03.034. <https://doi.org/10.1016/j.biocon.2018.03.034>.
- 13 Huang, S. K., L. Kuo, and K.-L. Chou, 2016: The applicability of marginal abatement cost approach:
14 A comprehensive review. *J. Clean. Prod.*, **127**, 59–71, doi:10.1016/J.JCLEPRO.2016.04.013.
15 <https://www.sciencedirect.com/science/article/pii/S0959652616302736> (Accessed April 14, 2019).
- 16 Huang, Y., and Coauthors, 2018: Impacts of species richness on productivity in a large-scale subtropical
17 forest experiment. *Science (80-.)*, **362**, 80–83, doi:10.1126/science.aat6405.
- 18 Hull, V., M.-N. Tuanniu, and J. Liu, 2015: Synthesis of human-nature feedbacks. *Ecol. Soc.*, **20**,
19 doi:1710.5751/es-07404-200317.
- 20 Humpenoder, F., and Coauthors, 2014: Investigating afforestation and bioenergy CCS as climate change
21 mitigation strategies. *Environ. Res. Lett.*, **9**, 064029, doi:064029 10.1088/1748-9326/9/6/064029.
22 [http://stacks.iop.org/1748-
23 9326/9/i=6/a=064029?key=crossref.5fa44a1462d2acebebaa002315d8e4a6](http://stacks.iop.org/1748-9326/9/i=6/a=064029?key=crossref.5fa44a1462d2acebebaa002315d8e4a6).
- 24 Humpenoeder, F., and Coauthors, 2018: Large-scale bioenergy production: how to resolve sustainability
25 trade-offs? *Environ. Res. Lett.*, **13**, doi:02401110.1088/1748-9326/aa9e3b.
- 26 Huq, S., and Coauthors, 2014: Adaptation Needs and Options. *Structure*, **14**.
- 27 Hussein, Z., T. Hertel, and A. Golub, 2013: Climate change mitigation policies and poverty in developing
28 countries. *Environ. Res. Lett.*, **8**, 035009, doi:10.1088/1748-9326/8/3/035009.
29 <http://stacks.iop.org/1748-9326/8/i=3/a=035009?key=crossref.98ac4b7e4a9ccd178df27ec1324bdf58>
30 (Accessed November 2, 2018).
- 31 Hussey, K., and J. Pittock, 2012: The Energy–Water Nexus: Managing the Links between Energy
32 and Water for a Sustainable Future. *Ecol. Soc.*, **17**, doi:10.5751/ES-04641-170131.
33 <https://www.ecologyandsociety.org/vol17/iss1/art31/>.
- 34 IFASTAT, 2018: Statistical Databases. <https://www.ifastat.org/>.
- 35 Iizumi, T., and N. Ramankutty, 2015: How do weather and climate influence cropping area and intensity?
36 *Glob. Food Sec.*, **4**, 46–50, doi:10.1016/j.gfs.2014.11.003.
- 37 IMF, 2018: *World Economic Outlook. World Economic Outlook Surveys, Oct 2018*. Washington D.C.,
38 [https://www.imf.org/en/Publications/WEO/Issues/2018/09/24/world-economic-outlook-october-
39 2018](https://www.imf.org/en/Publications/WEO/Issues/2018/09/24/world-economic-outlook-october-2018).
- 40 IPBES, 2016: *The methodological assessment report on scenarios and models of biodiversity and
41 ecosystem services*. S. Ferrier et al., Eds. Secretariat of the Intergovernmental Science-Policy
42 Platform on Biodiversity and Ecosystem Services, Bonn,.

- 1 —, 2018a: *IPBES Regional Assessment Report on Biodiversity and Ecosystem services from Europe*
2 *and Central Asia Biodiversity*. Bonn, Germany,
3 [https://www.ipbes.net/system/tdf/2018_eca_full_report_book_v5_pages_0.pdf?file=1&type=node&](https://www.ipbes.net/system/tdf/2018_eca_full_report_book_v5_pages_0.pdf?file=1&type=node&id=29180)
4 [id=29180](https://www.ipbes.net/system/tdf/2018_eca_full_report_book_v5_pages_0.pdf?file=1&type=node&id=29180).
- 5 —, 2018b: *The IPBES assessment report on land degradation and restoration*. Bonn, Germany, 744
6 pp.
7 [https://www.ipbes.net/system/tdf/2018_ldr_full_report_book_v4_pages.pdf?file=1&type=node&id=](https://www.ipbes.net/system/tdf/2018_ldr_full_report_book_v4_pages.pdf?file=1&type=node&id=29395)
8 [29395](https://www.ipbes.net/system/tdf/2018_ldr_full_report_book_v4_pages.pdf?file=1&type=node&id=29395).
- 9 —, 2018c: *IPBES Regional Assessment Report on Biodiversity and Ecosystem Services for Africa*.
10 Bonn, Germany, 492 pp.
11 [https://www.ipbes.net/system/tdf/africa_assessment_report_20181219_0.pdf?file=1&type=node&id](https://www.ipbes.net/system/tdf/africa_assessment_report_20181219_0.pdf?file=1&type=node&id=29243)
12 [=29243](https://www.ipbes.net/system/tdf/africa_assessment_report_20181219_0.pdf?file=1&type=node&id=29243).
- 13 —, 2018d: *The IPBES regional assessment report on biodiversity and ecosystem services for the*
14 *Americas*. Bonn, Germany, 656 pp.
15 [https://www.ipbes.net/system/tdf/2018_americas_full_report_book_v5_pages_0.pdf?file=1&type=n](https://www.ipbes.net/system/tdf/2018_americas_full_report_book_v5_pages_0.pdf?file=1&type=node&id=29404)
16 [ode&id=29404](https://www.ipbes.net/system/tdf/2018_americas_full_report_book_v5_pages_0.pdf?file=1&type=node&id=29404).
- 17 —, 2018e: *The IPBES regional assessment report on biodiversity and ecosystem services for Asia and*
18 *the Pacific*. Bonn, Germany, 612 pp.
19 [https://www.ipbes.net/system/tdf/2018_asia_pacific_full_report_book_v3_pages.pdf?file=1&type=n](https://www.ipbes.net/system/tdf/2018_asia_pacific_full_report_book_v3_pages.pdf?file=1&type=node&id=29507)
20 [ode&id=29507](https://www.ipbes.net/system/tdf/2018_asia_pacific_full_report_book_v3_pages.pdf?file=1&type=node&id=29507).
- 21 IPCC, 2018: *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of 1.5*
22 *°C above pre-industrial levels and related global greenhouse gas emission pathways, in the context*
23 *of strengthening the global response to the threat of climate change*.
24 <http://www.ipcc.ch/report/sr15/>.
- 25 Isbell, F., and Coauthors, 2017: Linking the influence and dependence of people on biodiversity across
26 scales. *Nature*, **546**, 65–72, doi:10.1038/nature22899. <http://dx.doi.org/10.1038/nature22899>.
- 27 Iwata, Y., T. Miyamoto, K. Kameyama, and M. Nishiya, 2017: Effect of sensor installation on the
28 accurate measurement of soil water content. *Eur. J. Soil Sci.*, **68**, 817–828, doi:10.1111/ejss.12493.
- 29 Jadin, I., P. Meyfroidt, and E. F. Lambin, 2016: International trade, and land use intensification and
30 spatial reorganization explain Costa Rica’s forest transition. *Environ. Res. Lett.*, **11**, 035005,
31 doi:10.1088/1748-9326/11/3/035005. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=3/a=035005?key=crossref.19829b837de35e3e24487f52b50343fe)
32 [9326/11/i=3/a=035005?key=crossref.19829b837de35e3e24487f52b50343fe](http://stacks.iop.org/1748-9326/11/i=3/a=035005?key=crossref.19829b837de35e3e24487f52b50343fe).
- 33 Jalava, M., M. Kummu, M. Porkka, S. Siebert, and O. Varis, 2014: Diet change—a solution to reduce
34 water use? *Environ. Res. Lett.*, **9**, 74016. <http://stacks.iop.org/1748-9326/9/i=7/a=074016>.
- 35 EL Jarroudi, M., and Coauthors, 2015: Economics of a decision–support system for managing the main
36 fungal diseases of winter wheat in the Grand-Duchy of Luxembourg. *F. Crop. Res.*, **172**, 32–41,
37 doi:10.1016/J.FCR.2014.11.012.
38 <https://www.sciencedirect.com/science/article/pii/S0378429014003281?via%3Dihub> (Accessed
39 May 21, 2018).
- 40 Jiang, L., and B. C. O’Neill, 2017: Global urbanization projections for the Shared Socioeconomic
41 Pathways. *Glob. Environ. Chang.*, **42**, 193–199, doi:10.1016/J.GLOENVCHA.2015.03.008.
42 <https://www.sciencedirect.com/science/article/pii/S0959378015000394?via%3Dihub> (Accessed
43 April 23, 2018).

- 1 de Jong, R., M. E. Schaepman, R. Furrer, S. de Bruin, and P. H. Verburg, 2013: Spatial relationship
2 between climatologies and changes in global vegetation activity. *Glob. Chang. Biol.*,
3 doi:10.1111/gcb.12193.
- 4 Joshi, A. K., P. Pant, P. Kumar, A. Giriraj, and P. K. Joshi, 2011: National Forest Policy in India: Critique
5 of Targets and Implementation. *Small-scale For.*, **10**, 83–96, doi:10.1007/s11842-010-9133-z.
6 <http://link.springer.com/10.1007/s11842-010-9133-z> (Accessed August 16, 2018).
- 7 Juhl, H. J., and M. B. Jensen, 2014: Relative price changes as a tool to stimulate more healthy food
8 choices – A Danish household panel study. *Food Policy*, **46**, 178–182,
9 doi:10.1016/J.FOODPOL.2014.03.008.
10 <https://www.sciencedirect.com/science/article/pii/S0306919214000487> (Accessed May 24, 2018).
- 11 Kaijser, A., and A. Kronsell, 2014: Climate change through the lens of intersectionality. *Env. Polit.*, **23**,
12 417–433, doi:10.1080/09644016.2013.835203.
13 <http://www.tandfonline.com/doi/abs/10.1080/09644016.2013.835203> (Accessed April 17, 2019).
- 14 Kanter, D. R., and Coauthors, 2016: Evaluating agricultural trade-offs in the age of sustainable
15 development ☆. *AGSY*, doi:10.1016/j.agsy.2016.09.010.
16 <http://dx.doi.org/10.1016/j.agsy.2016.09.010>.
- 17 Kastner, T., M. J. I. Rivas, W. Koch, and S. Nonhebel, 2012: Global changes in diets and the
18 consequences for land requirements for food. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.1117054109.
- 19 Kastner, T., K. H. Erb, and H. Haberl, 2014: Rapid growth in agricultural trade: Effects on global area
20 efficiency and the role of management. *Environ. Res. Lett.*, **9**, doi:10.1088/1748-9326/9/3/034015.
- 21 Kauffman, J. B., H. Hernandez Trejo, M. del Carmen Jesus Garcia, C. Heider, and W. M. Contreras,
22 2016: Carbon stocks of mangroves and losses arising from their conversion to cattle pastures in the
23 Pantanos de Centla, Mexico. *Wetl. Ecol. Manag.*, **24**, 203–216, doi:10.1007/s11273-015-9453-z.
24 <https://doi.org/10.1007/s11273-015-9453-z>.
- 25 De Kauwe, M. G., T. F. Keenan, B. E. Medlyn, I. C. Prentice, and C. Terrer, 2016: Satellite based
26 estimates underestimate the effect of CO2 fertilization on net primary productivity. *Nat. Clim.*
27 *Chang.*, doi:10.1038/nclimate3105.
- 28 Keenan, R. J., 2015: Climate Change Impacts and Adaptation in Forest Management: A Review. *Ann.*
29 *For. Sci.*, **72**, 145–167, doi:10.1007/s13595-014-0446-5.
- 30 Keenan, R. J., G. A. Reams, F. Achard, J. V de Freitas, A. Grainger, and E. Lindquist, 2015: Dynamics of
31 global forest area: Results from the FAO Global Forest Resources Assessment 2015. *For. Ecol.*
32 *Manage.*, **352**, 9–20, doi:10.1016/j.foreco.2015.06.014.
33 <https://www.sciencedirect.com/science/article/pii/S0378112715003400> (Accessed November 22,
34 2017).
- 35 Kelley, D. I., I. C. Prentice, S. P. Harrison, H. Wang, M. Simard, J. B. Fisher, and K. O. Willis, 2013: A
36 comprehensive benchmarking system for evaluating global vegetation models. *Biogeosciences*, **10**,
37 3313–3340, doi:10.5194/bg-10-3313-2013. <http://www.biogeosciences.net/10/3313/2013/>.
- 38 Kesicki, F., 2013: What are the key drivers of MAC curves? A partial-equilibrium modelling approach for
39 the UK. *Energy Policy*, **58**, 142–151, doi:10.1016/J.ENPOL.2013.02.043.
40 <https://www.sciencedirect.com/science/article/pii/S0301421513001493> (Accessed April 3, 2019).
- 41 Kibler, K. M., D. Reinhart, C. Hawkins, A. M. Motlagh, and J. Wright, 2018: Food waste and the food-
42 energy-water nexus: A review of food waste management alternatives. *Waste Manag.*, **74**, 52–62,
43 doi:10.1016/J.WASMAN.2018.01.014.
44 <https://www.sciencedirect.com/science/article/pii/S0956053X18300151> (Accessed May 24, 2018).

- 1 Kimball, B. A., 2016: Crop responses to elevated CO₂ and interactions with H₂O, N, and temperature.
2 *Curr. Opin. Plant Biol.*, **31**, 36–43, doi:<https://doi.org/10.1016/j.pbi.2016.03.006>.
3 <http://www.sciencedirect.com/science/article/pii/S1369526616300334>.
- 4 Kindermann, G., I. McCallum, S. Fritz, and M. Obersteiner, 2008: A global forest growing stock, biomass
5 and carbon map based on FAO statistics. *Silva Fenn.*, **42**, doi:10.14214/sf.244.
6 <http://www.silvafennica.fi/article/244> (Accessed April 3, 2019).
- 7 Klein, J. A., K. A. Hopping, E. T. Yeh, Y. Nyima, R. B. Boone, and K. A. Galvin, 2014: Unexpected
8 climate impacts on the Tibetan Plateau: Local and scientific knowledge in findings of delayed
9 summer. *Glob. Environ. Chang.*, **28**, 141–152, doi:10.1016/J.GLOENVCHA.2014.03.007.
10 <https://www.sciencedirect.com/science/article/pii/S0959378014000557> (Accessed May 24, 2018).
- 11 Klein, R. J. T., S. Huq, F. Denton, T. E. Downing, R. G. Richels, J. B. Robinson, and F. L. Toth, 2007:
12 Inter-relationships between adaptation and mitigation. Climate Change 2007: Impacts, Adaptation
13 and Vulnerability. *Contribution of Working Group II to the Fourth Assessment Report of the*
14 *Intergovernmental Panel on Climate Change*, Cambridge University Press, Cambridge, UK, 745–
15 777.
- 16 Kok, K., I. Bärlund, M. Flörke, I. Holman, M. Gramberger, J. Sendzimir, B. Stuch, and K. Zellmer, 2014:
17 European participatory scenario development: strengthening the link between stories and models.
18 *Clim. Change*, **128**, 187–200, doi:10.1007/s10584-014-1143-y.
- 19 Kok, M. T. J., and Coauthors, 2018: Pathways for agriculture and forestry to contribute to terrestrial
20 biodiversity conservation: A global scenario-study. *Biol. Conserv.*, **221**, 137–150,
21 doi:10.1016/j.biocon.2018.03.003.
- 22 Kolby Smith, W., S. C. Reed, C. C. Cleveland, A. P. Ballantyne, W. R. L. Anderegg, W. R. Wieder, Y. Y.
23 Liu, and S. W. Running, 2015: Large divergence of satellite and Earth system model estimates of
24 global terrestrial CO₂ fertilization. *Nat. Clim. Chang.*, doi:10.1038/nclimate2879.
- 25 Konar, M., J. J. Reimer, Z. Hussein, and N. Hanasaki, 2016: The water footprint of staple crop trade
26 under climate and policy scenarios. *Environ. Res. Lett.*, **11**, 035006, doi:10.1088/1748-
27 9326/11/3/035006. [http://stacks.iop.org/1748-
28 9326/11/i=3/a=035006?key=crossref.76328e79ac5ddc32bd0fd20ad2cb8f3f](http://stacks.iop.org/1748-9326/11/i=3/a=035006?key=crossref.76328e79ac5ddc32bd0fd20ad2cb8f3f) (Accessed March 22,
29 2019).
- 30 Kongsager, R., and E. Corbera, 2015: Linking Mitigation and Adaptation in Carbon Forestry Projects:
31 Evidence from Belize. *World Dev.*, **76**, 132–146, doi:10.1016/J.WORLDDEV.2015.07.003.
32 <https://www.sciencedirect.com/science/article/pii/S0305750X1500162X> (Accessed April 3, 2019).
- 33 ———, B. Locatelli, and F. Chazarin, 2016: Addressing Climate Change Mitigation and Adaptation
34 Together: A Global Assessment of Agriculture and Forestry Projects. *Environ. Manage.*, **57**, 271–
35 282, doi:<http://dx.doi.org/10.1007/s00267-015-0605-y>.
36 <https://search.proquest.com/docview/1756407729?accountid=10673>.
- 37 Kostyanovsky, K. I., D. R. Huggins, C. O. Stockle, S. Waldo, and B. Lamb, 2018: Developing a flow
38 through chamber system for automated measurements of soil N₂O and CO₂ emissions. *Meas. J. Int.*
39 *Meas. Confed.*, **113**, 172–180, doi:10.1016/j.measurement.2017.05.040.
40 <http://dx.doi.org/10.1016/j.measurement.2017.05.040>.
- 41 Koutroulis, A. G., 2019: Dryland changes under different levels of global warming. *Sci. Total Environ.*,
42 **655**, 482–511, doi:10.1016/J.SCITOTENV.2018.11.215.
43 <https://www.sciencedirect.com/science/article/pii/S0048969718345716?via%3Dihub> (Accessed
44 April 10, 2019).

- 1 Krause, A., and Coauthors, 2017: Global consequences of afforestation and bioenergy cultivation on
2 ecosystem service indicators. *Biogeosciences*, **2017**, 4829–4850, doi:10.5194/bg-2017-160.
3 <https://www.biogeosciences-discuss.net/bg-2017-160/>.
- 4 Krausmann, F., and E. Langthaler, 2019: Food regimes and their trade links: A socio-ecological
5 perspective. *Ecol. Econ.*, **160**, 87–95, doi:10.1016/J.ECOLECON.2019.02.011.
6 <https://www.sciencedirect.com/science/article/pii/S0921800918317300> (Accessed March 22, 2019).
- 7 ———, and Coauthors, 2013: Global human appropriation of net primary production doubled in the 20th
8 century. *Proc. Natl. Acad. Sci. U. S. A.*, **110**, 10324–10329, doi:10.1073/pnas.1211349110.
- 9 Kraxner, F., and Coauthors, 2013: Global bioenergy scenarios - Future forest development, land-use
10 implications, and trade-offs. *Biomass and Bioenergy*, **57**, 86–96,
11 doi:10.1016/j.biombioe.2013.02.003.
- 12 Kreidenweis, U., F. Humpenöder, M. Stevanovic, B. L. Bodirsky, E. Kriegler, H. Lotze-Campen, and A.
13 Popp, 2016: Afforestation to mitigate climate change: impacts on food prices under consideration of
14 albedo effects. *Environ. Res. Lett.*, **11**, 1–12, doi:doi:10.1088/1748-9326/11/8/085001.
- 15 ———, ———, L. Kehoe, T. Kuemmerle, B. L. Bodirsky, H. Lotze-Campen, and A. Popp, 2018: Pasture
16 intensification is insufficient to relieve pressure on conservation priority areas in open agricultural
17 markets. *Glob. Chang. Biol.*, **24**, 3199–3213, doi:10.1111/gcb.14272.
- 18 Kummu, M., H. de Moel, M. Porkka, S. Siebert, O. Varis, and P. J. Ward, 2012: Lost food, wasted
19 resources: Global food supply chain losses and their impacts on freshwater, cropland, and fertiliser
20 use. *Sci. Total Environ.*, **438**, 477–489, doi:10.1016/J.SCITOTENV.2012.08.092.
21 <https://www.sciencedirect.com/science/article/pii/S0048969712011862> (Accessed April 3, 2019).
- 22 Kunreuther, H., 2015: The Role of Insurance in Reducing Losses from Extreme Events: The Need for
23 Public–Private Partnerships. *Geneva Pap. Risk Insur. - Issues Pract.*, **40**, 741–762,
24 doi:10.1057/gpp.2015.14. <https://doi.org/10.1057/gpp.2015.14>.
- 25 Lacaze, R., and Coauthors, 2015: Operational 333m biophysical products of the copernicus global land
26 service for agriculture monitoring. *Int. Arch. Photogramm. Remote Sens. Spat. Inf. Sci. - ISPRS*
27 *Arch.*, **40**, 53–56, doi:10.5194/isprsarchives-XL-7-W3-53-2015.
- 28 Laestadius, L., S. Maginnis, S. Minnemeyer, P. Potapov, C. Saint-Laurent, and N. Sizer, 2011: Mapping
29 opportunities for forest landscape restoration. *Unasylva*, **62**, 47–48.
- 30 Lal, R., 2009: Soils and world food security. *Soil and Tillage Research*.
- 31 Lal, R., 2015: Restoring Soil Quality to Mitigate Soil Degradation. *Sustainability*, **7**, 5875.
32 <http://www.mdpi.com/2071-1050/7/5/5875>.
- 33 Lamb, D., 2018: Undertaking large-scale forest restoration to generate ecosystem services. *Restor. Ecol.*,
34 **26**, 657–666, doi:10.1111/rec.12706.
- 35 Lambin, E. F., 2012: *Global land availability: Malthus versus Ricardo*. Elsevier,
36 <https://www.sciencedirect.com/science/article/pii/S2211912412000235> (Accessed December 4,
37 2012).
- 38 Lambin, E. F., and P. Meyfroidt, 2011: Global land use change, economic globalization, and the looming
39 land scarcity. *Proc Natl Acad Sci U S A*, **108**, 3465–3472, doi:10.1073/pnas.1100480108.
- 40 Lambin, E. F., and P. Meyfroidt, 2014: Trends in Global Land-Use Competition. *Rethinking Global Land*
41 *Use in an Urban Era*, Vol. 14 of, The MIT Press, 11–22.
- 42 Land Matrix, 2018: Land Matrix Global Observatory. <http://www.landmatrix.org>.

- 1 Lapola, D. M., R. Schaldach, J. Alcamo, A. Bondeau, J. Koch, C. Koelking, and J. A. Priess, 2010:
2 Indirect land-use changes can overcome carbon savings from biofuels in Brazil. *Proc. Natl. Acad.
3 Sci. U. S. A.*, **107**, 3388–3393, doi:10.1073/pnas.0907318107.
- 4 Lassaletta, L., G. Billen, B. Grizzetti, J. Anglade, and J. Garnier, 2014: 50 year trends in nitrogen use
5 efficiency of world cropping systems: The relationship between yield and nitrogen input to
6 cropland. *Environ. Res. Lett.*, doi:10.1088/1748-9326/9/10/105011.
- 7 ———, ———, J. Garnier, L. Bouwman, E. Velazquez, N. D. Mueller, and J. S. Gerber, 2016: Nitrogen use
8 in the global food system: past trends and future trajectories of agronomic performance, pollution,
9 trade, and dietary demand. *Environ. Res. Lett.*, **11**, 095007, doi:10.1088/1748-9326/11/9/095007.
10 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=9/a=095007?key=crossref.0796ce04da0253665bff8dd7a28b13e4)
11 [9326/11/i=9/a=095007?key=crossref.0796ce04da0253665bff8dd7a28b13e4](http://stacks.iop.org/1748-9326/11/i=9/a=095007?key=crossref.0796ce04da0253665bff8dd7a28b13e4) (Accessed March 22,
12 2019).
- 13 Laurance, W. F., 2007: Forests and floods. *Nature*, **449**, 409–410, doi:10.1111/j.1365-2486.200701446.x.
- 14 Laurance, W. F., J. Sayer, and K. G. Cassman, 2014: Agricultural expansion and its impacts on tropical
15 nature. *Trends Ecol. Evol.*, **29**, 107–116, doi:10.1016/J.TREE.2013.12.001.
16 <https://www.sciencedirect.com/science/article/pii/S0169534713002929> (Accessed April 23, 2018).
- 17 Le, H. D., C. Smith, J. Herbohn, and S. Harrison, 2012: More than just trees: Assessing reforestation
18 success in tropical developing countries. *J. Rural Stud.*, **28**, 5–19,
19 doi:10.1016/j.jrurstud.2011.07.006.
- 20 Le, Q. B., E. Nkonya, and A. Mirzabaev, 2016: Biomass Productivity-Based Mapping of Global Land
21 Degradation Hotspots. *Economics of Land Degradation and Improvement – A Global Assessment
22 for Sustainable Development*, E. Nkonya, A. Mirzabaev, and J. Von Braun, Eds., Springer
23 International Publishing, Cham, 55–84 https://doi.org/10.1007/978-3-319-19168-3_4.
- 24 LE, T. T. H., 2016: EFFECTS OF CLIMATE CHANGE ON RICE YIELD AND RICE MARKET IN
25 VIETNAM. *J. Agric. Appl. Econ.*, **48**, 366–382, doi:10.1017/aae.2016.21.
26 https://www.cambridge.org/core/product/identifier/S1074070816000213/type/journal_article
27 (Accessed January 19, 2018).
- 28 Lean, J. L., 2018: Observation-based detection and attribution of 21st century climate change. *Wiley
29 Interdiscip. Rev. Chang.*, **9**, doi:UNSP e51110.1002/wcc.511.
- 30 Lebel, L., J. M. Anderies, B. Campbell, C. Folke, S. Hatfield-Dodds, T. P. Hughes, and J. Wilson, 2006:
31 Governance and the Capacity to Manage Resilience in Regional Social-Ecological Systems. *Ecol.
32 Soc.*, **11**, 19. <http://www.ecologyandsociety.org/vol11/iss1/art19/>.
- 33 Lee, J., C. H. Lim, G. S. Kim, A. Markandya, S. Chowdhury, S. J. Kim, W. K. Lee, and Y. Son, 2018:
34 Economic viability of the national-scale forestation program: The case of success in the Republic of
35 Korea. *Ecosyst. Serv.*, **29**, 40–46, doi:10.1016/j.ecoser.2017.11.001.
36 <https://doi.org/10.1016/j.ecoser.2017.11.001>.
- 37 Lee, X., and Coauthors, 2011: Observed increase in local cooling effect of deforestation at higher
38 latitudes. *Nature*, **479**, 384–387,
39 doi:[http://www.nature.com/nature/journal/v479/n7373/abs/nature10588.html#supplementary-](http://www.nature.com/nature/journal/v479/n7373/abs/nature10588.html#supplementary-information)
40 [information. http://dx.doi.org/10.1038/nature10588](http://dx.doi.org/10.1038/nature10588).
- 41 Lees, K. J., T. Quafe, R. R. E. Artz, M. Khomik, and J. M. Clark, 2018: Potential for using remote
42 sensing to estimate carbon fluxes across northern peatlands – A review. *Sci. Total Environ.*, **615**,
43 857–874, doi:10.1016/j.scitotenv.2017.09.103. <https://doi.org/10.1016/j.scitotenv.2017.09.103>.

- 1 Lehmann, C. E. R., and C. L. Parr, 2016: Tropical grassy biomes: linking ecology, human use and
2 conservation. *Philos. Trans. R. Soc. B-Biological Sci.*, **371**, 20160329, doi:20160329
3 10.1098/rstb.2016.0329.
- 4 Lehmann, N., S. Briner, and R. Finger, 2013: The impact of climate and price risks on agricultural land
5 use and crop management decisions. *Land use policy*, **35**, 119–130,
6 doi:10.1016/J.LANDUSEPOL.2013.05.008.
7 <https://www.sciencedirect.com/science/article/pii/S0264837713000902> (Accessed January 19,
8 2018).
- 9 Lempert, R., N. Nakicenovic, D. Sarewitz, and M. Schlesinger, 2004: Characterizing Climate-Change
10 Uncertainties for Decision-Makers. An Editorial Essay. *Clim. Change*, **65**, 1–9,
11 doi:10.1023/B:CLIM.0000037561.75281.b3.
12 <http://link.springer.com/10.1023/B:CLIM.0000037561.75281.b3>.
- 13 Lenton, T. M., 2014: The Global Potential for Carbon Dioxide Removal. *Geoengineering of the Climate*
14 *System*, R.E. Hester and R.M. Harrison, Eds., Royal Society of Chemistry, 52–79.
- 15 Lesk, C., P. Rowhani, and N. Ramankutty, 2016: Influence of extreme weather disasters on global crop
16 production. *Nature*, **529**, 84. <http://dx.doi.org/10.1038/nature16467>.
- 17 Li, D., S. Niu, and Y. Luo, 2012: Global patterns of the dynamics of soil carbon and nitrogen stocks
18 following afforestation: A meta-analysis. *New Phytol.*, **195**, 172–181, doi:10.1111/j.1469-
19 8137.2012.04150.x. [http://www.scopus.com/inward/record.url?eid=2-s2.0-
20 84861480327&partnerID=40&md5=c9b5ed6ec5d9b065030cbc3f85147047](http://www.scopus.com/inward/record.url?eid=2-s2.0-84861480327&partnerID=40&md5=c9b5ed6ec5d9b065030cbc3f85147047).
- 21 Li, S., M. Xu, and B. Sun, 2014: Long-term hydrological response to reforestation in a large watershed in
22 southeastern China. *Hydrol. Process.*, **28**, 5573–5582, doi:10.1002/hyp.10018.
- 23 Li, W., P. Ciais, N. MacBean, S. Peng, P. Defourny, and S. Bontemps, 2016: Major forest changes and
24 land cover transitions based on plant functional types derived from the ESA CCI Land Cover
25 product. *Int. J. Appl. Earth Obs. Geoinf.*, **47**, 30–39, doi:10.1016/J.JAG.2015.12.006.
26 <https://www.sciencedirect.com/science/article/pii/S0303243415300714> (Accessed April 7, 2019).
- 27 —, and Coauthors, 2017: Land-use and land-cover change carbon emissions between 1901 and 2012
28 constrained by biomass observations. *Biogeosciences*, **145194**, 5053–5067, doi:10.5194/bg-14-
29 5053-2017.
- 30 Limpens, J., F. Berendse, C. Blodau, J. G. Canadell, C. Freeman, J. Holden, N. Roulet, and H. Rydin,
31 2008: Peatlands and the carbon cycle: from local processes to global implications – a synthesis.
32 *Biogeosciences*, **5**, 1475–1491.
- 33 Lin, M., and P. Huybers, 2012: Reckoning wheat yield trends. *Environ. Res. Lett.*, **7**, 24016.
34 <http://stacks.iop.org/1748-9326/7/i=2/a=024016>.
- 35 Lindenmayer, D. B., and R. J. Hobbs, 2004: Fauna conservation in Australian plantation forests – a
36 review. *Biol. Conserv.*, **119**, 151–168, doi:10.1016/J.BIOCON.2003.10.028.
37 <https://www.sciencedirect.com/science/article/pii/S0006320703004439> (Accessed August 17,
38 2018).
- 39 Linnerooth-Bayer, J., and R. Mechler, 2006: Insurance for assisting adaptation to climate change in
40 developing countries: A proposed strategy. *Clim. Policy*, **6**, 621–636,
41 doi:10.1080/14693062.2006.9685628.
- 42 Lipper, L., and Coauthors, 2014: Climate-smart agriculture for food security. *Nat. Clim. Chang.*, **4**, 1068–
43 1072, doi:10.1038/nclimate2437. <http://www.nature.com/articles/nclimate2437> (Accessed April 24,
44 2018).

- 1 Liu, J., and Coauthors, 2013: Framing Sustainability in a Telecoupled World. *Ecol. Soc.*, **2**,
2 doi:<http://dx.doi.org/10.5751/ES-05873-180226>.
3 http://curis.ku.dk/ws/files/46234860/2013_liu_et_al_ES_2013_5873.pdf (Accessed September 16,
4 2014).
- 5 Lobell, D. B., W. Schlenker, and J. Costa-Roberts, 2011: Climate Trends and Global Crop Production
6 Since 1980. *Science (80-.)*, **333**, 616–620, doi:10.1126/science.1204531.
7 <http://www.sciencemag.org/content/333/6042/616.abstract>.
- 8 ———, A. Sibley, and J. Ivan Ortiz-Monasterio, 2012: Extreme heat effects on wheat senescence in India.
9 *Nat. Clim. Chang.*, **2**, 186–189,
10 doi:[http://www.nature.com/nclimate/journal/v2/n3/abs/nclimate1356.html#supplementary-](http://www.nature.com/nclimate/journal/v2/n3/abs/nclimate1356.html#supplementary-information)
11 [information](http://dx.doi.org/10.1038/nclimate1356). <http://dx.doi.org/10.1038/nclimate1356>.
- 12 ———, C. B. Uris Lantz, and W. H. Thomas, 2013: Climate adaptation as mitigation: the case of
13 agricultural investments. *Environ. Res. Lett.*, **8**, 15012. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/8/i=1/a=015012)
14 [9326/8/i=1/a=015012](http://stacks.iop.org/1748-9326/8/i=1/a=015012).
- 15 Locatelli, B., V. Evans, A. Wardell, A. Andrade, and R. Vignola, 2011: Forests and Climate Change in
16 Latin America: Linking Adaptation and Mitigation. *Forests*, **2**, doi:10.3390/f2010431.
- 17 ———, and Coauthors, 2015a: Tropical reforestation and climate change: beyond carbon. *Restor. Ecol.*, **23**,
18 337–343, doi:10.1111/rec.12209. <http://doi.wiley.com/10.1111/rec.12209> (Accessed September 19,
19 2018).
- 20 ———, C. Pavageau, E. Pramova, and M. Di Gregorio, 2015b: Integrating climate change mitigation and
21 adaptation in agriculture and forestry: opportunities and trade-offs. *Wiley Interdiscip. Rev. Clim.*
22 *Chang.*, **6**, n/a-n/a, doi:10.1002/wcc.357. <http://doi.wiley.com/10.1002/wcc.357> (Accessed April 3,
23 2019).
- 24 Loch, A., S. Wheeler, H. Bjornlund, B. S. J. Edwards, A. Zuo, and M. Shanahan, 2013: *The role of water*
25 *markets in climate change adaptation*. National Climate Change Adaptation Research Facility Gold
26 Coast,.
- 27 Loladze, I., 2014: Hidden shift of the ionome of plants exposed to elevated CO(2) depletes minerals at the
28 base of human nutrition. *Elife*, **3**, e02245, doi:10.7554/eLife.02245.
29 <http://www.ncbi.nlm.nih.gov/pmc/articles/PMC4034684/>.
- 30 Lorenz, K., and R. Lal, 2014: Biochar application to soil for climate change mitigation by soil organic
31 carbon sequestration. *J. Plant Nutr. Soil Sci.*, **177**, 651–670, doi:10.1002/jpln.201400058.
- 32 Luedeling, E., and E. Shepherd, 2016: Decision-Focused Agricultural Research. *Solutions*, **7**, 46–54.
- 33 Luo, Y. Q., and Coauthors, 2012: A framework of benchmarking land models. *Biogeosciences*, **10**, 3857–
34 3874, doi:10.5194/bgd-9-1899-2012.
- 35 Luyssaert, S., and Coauthors, 2014: Land management and land-cover change have impacts of similar
36 magnitude on surface temperature. *Nat. Clim. Chang.*, **4**, 389–393, doi:10.1038/nclimate2196.
37 <http://www.nature.com/nclimate/journal/v4/n5/full/nclimate2196.html> (Accessed July 15, 2014).
- 38 MacDicken, K. G., 2015: Global Forest Resources Assessment 2015: What, why and how? *For. Ecol.*
39 *Manage.*, **352**, 3–8, doi:10.1016/j.foreco.2015.02.006.
40 <http://dx.doi.org/10.1016/j.foreco.2015.02.006>.
- 41 MacDicken, K. G., P. Sola, J. E. Hall, C. Sabogal, M. Tadoum, and C. de Wasseige, 2015: Global
42 progress toward sustainable forest management. *For. Ecol. Manage.*, **352**, 47–56,
43 doi:<https://doi.org/10.1016/j.foreco.2015.02.005>.
44 <http://www.sciencedirect.com/science/article/pii/S0378112715000560>.

- 1 Mace, G. M., K. Norris, and A. H. Fitter, 2012: Biodiversity and ecosystem services: A multilayered
2 relationship. *Trends Ecol. Evol.*, **27**, 19–25, doi:10.1016/j.tree.2011.08.006.
- 3 Maestre, F. T., and Coauthors, 2012: Plant Species Richness and Ecosystem Multifunctionality in Global
4 Drylands. *Science (80-.)*, **335**, 214–218, doi:10.1126/science.1215442.
- 5 ———, and Coauthors, 2016: Structure and Functioning of Dryland Ecosystems in a Changing World.
6 *Annual Review of Ecology, Evolution, and Systematics, Vol 47*, D.J. Futuyma, Ed., Vol. 47 of, 215–
7 237.
- 8 Maier, H. R., J. H. A. Guillaume, H. van Delden, G. A. Riddell, M. Haasnoot, and J. H. Kwakkel, 2016:
9 An uncertain future, deep uncertainty, scenarios, robustness and adaptation: How do they fit
10 together? *Environ. Model. Softw.*, **81**, 154–164, doi:10.1016/j.envsoft.2016.03.014.
- 11 Malkamäki, A., D. D’Amato, N. J. Hogarth, M. Kanninen, R. Pirard, A. Toppinen, and W. Zhou, 2018: A
12 systematic review of the socio-economic impacts of large-scale tree plantations, worldwide. *Glob.*
13 *Environ. Chang.*, **53**, 90–103, doi:10.1016/j.gloenvcha.2018.09.001.
14 <https://doi.org/10.1016/j.gloenvcha.2018.09.001>.
- 15 Malone, R. W., and Coauthors, 2014: Cover crops in the upper midwestern United States: Simulated
16 effect on nitrate leaching with artificial drainage. *J. Soil Water Conserv.*, **69**, 292–305,
17 doi:10.2489/jswc.69.4.292. <http://www.jswconline.org/cgi/doi/10.2489/jswc.69.4.292> (Accessed
18 April 3, 2019).
- 19 Marchand, P., and Coauthors, 2016: Reserves and trade jointly determine exposure to food supply shocks.
20 *Environ. Res. Lett.*, **11**, 095009, doi:10.1088/1748-9326/11/9/095009. [http://stacks.iop.org/1748-
21 9326/11/i=9/a=095009?key=crossref.15f8547c7864d2f76a79429020178381](http://stacks.iop.org/1748-9326/11/i=9/a=095009?key=crossref.15f8547c7864d2f76a79429020178381).
- 22 Marques, A., and Coauthors, 2019: Increasing impacts of land use on biodiversity and carbon
23 sequestration driven by population and economic growth. *Nat. Ecol. Evol.*, **1**, doi:10.1038/s41559-
24 019-0824-3. <http://www.nature.com/articles/s41559-019-0824-3> (Accessed March 21, 2019).
- 25 Martellozzo, F., N. Ramankutty, R. J. Hall, D. T. Price, B. Purdy, and M. A. Friedl, 2015: Urbanization
26 and the loss of prime farmland: a case study in the Calgary–Edmonton corridor of Alberta. *Reg.*
27 *Environ. Chang.*, **15**, 881–893, doi:10.1007/s10113-014-0658-0.
28 <http://link.springer.com/10.1007/s10113-014-0658-0> (Accessed April 24, 2018).
- 29 Martin-Guay, M. O., A. Paquette, J. Dupras, and D. Rivest, 2018: The new Green Revolution: Sustainable
30 intensification of agriculture by intercropping. *Sci. Total Environ.*, **615**, 767–772,
31 doi:10.1016/j.scitotenv.2017.10.024.
- 32 Mastrandrea, M. D., K. J. Mach, G.-K. Plattner, O. Edenhofer, T. F. Stocker, C. B. Field, K. L. Ebi, and
33 P. R. Matschoss, 2011: The IPCC AR5 guidance note on consistent treatment of uncertainties: a
34 common approach across the working groups. *Clim. Change*, **108**, 675, doi:10.1007/s10584-011-
35 0178-6. <https://doi.org/10.1007/s10584-011-0178-6>.
- 36 Mateos, E., J. M. Edeso, and L. Ormaetxea, 2017: Soil Erosion and Forests Biomass as Energy Resource
37 in the Basin of the Oka River in Biscay ., *Forests*, **8**, 1–20, doi:10.3390/f8070258.
- 38 Mathews, J. A., 2017: Global trade and promotion of cleantech industry: a post-Paris agenda. *Clim.*
39 *Policy*, **17**, 102–110, doi:10.1080/14693062.2016.1215286.
40 <https://doi.org/10.1080/14693062.2016.1215286>.
- 41 Maxwell, S. L., R. A. Fuller, T. M. Brooks, and J. E. M. Watson, 2016: Biodiversity: The ravages of
42 guns, nets and bulldozers. *Nature*, doi:10.1038/536143a.
- 43 McDonnell, S., 2017: *Urban Land Grabbing by Political Elites: Exploring the Political Economy of Land
44 and the Challenges of Regulation. In: Kastom, property and ideology: land transformations in*

- 1 *Melanesia*. C.F. McDonnell, S., M. G. Allen, C. Filer (editors). McDonnell, S., M. G. Allen, Ed.
2 American National University Press,.
- 3 Medek, Danielle E., Joel Schwartz, S. S. M., 2017: "Estimated effects of future atmospheric CO2
4 concentrations on protein intake and the risk of protein deficiency by country and region. *Env. Heal.*
5 *Perspect*, **125**, 087002. https://ehp.niehs.nih.gov/ehp41/?utm_source=rss.
- 6 Mello, D., and M. Schmink, 2017: Amazon entrepreneurs: Women's economic empowerment and the
7 potential for more sustainable land use practices. *Womens. Stud. Int. Forum*, **65**, 28–36,
8 doi:10.1016/J.WSIF.2016.11.008.
9 <https://www.sciencedirect.com/science/article/pii/S027753951530176X> (Accessed April 3, 2019).
- 10 Messerli, P., M. Giger, M. B. Dwyer, T. Breu, and S. Eckert, 2014: The geography of large-scale land
11 acquisitions: Analysing socio-ecological patterns of target contexts in the global South. *Appl.*
12 *Geogr.*, **53**, 449–459, doi:10.1016/j.apgeog.2014.07.005.
13 <http://www.sciencedirect.com/science/article/pii/S0143622814001611>.
- 14 Meyfroidt, P., 2018: Trade-offs between environment and livelihoods: Bridging the global land use and
15 food security discussions. *Glob. Food Sec.*, **16**, 9–16, doi:10.1016/J.GFS.2017.08.001.
16 <https://www.sciencedirect.com/science/article/pii/S2211912416301067> (Accessed April 23, 2018).
- 17 Millar, R. J., and Coauthors, 2017: Emission budgets and pathways consistent with limiting warming to
18 1.5 C. *Nat. Geosci.*, doi:DOI: 10.1038/NGEO3031.
- 19 Mirzabaev, A., E. Nkonya, and J. von Braun, 2015: *Economics of sustainable land management*. Elsevier,
20 <https://www.sciencedirect.com/science/article/pii/S1877343515000688> (Accessed April 25, 2018).
- 21 Mistry, J., and A. Berardi, 2016: Bridging indigenous and scientific knowledge. *Science (80-.)*, **352**,
22 1274 LP-1275. <http://science.sciencemag.org/content/352/6291/1274.abstract>.
- 23 Miyamoto, A., M. Sano, H. Tanaka, and K. Niiyama, 2011: Changes in forest resource utilization and
24 forest landscapes in the southern Abukuma Mountains, Japan during the twentieth century. *J. For.*
25 *Res.*, **16**, 87–97, doi:10.1007/s10310-010-0213-x.
- 26 Molina, A., G. Govers, V. Vanacker, and J. Poesen, 2007: Author ' s personal copy Runoff generation in
27 a degraded Andean ecosystem : Interaction of vegetation cover and land use. *Catena*, **71**, 357–370,
28 doi:10.1016/j.catena.2007.04.002.
- 29 Moore, F. C., and D. B. Lobell, 2015: The fingerprint of climate trends on European crop yields. *Proc.*
30 *Natl. Acad. Sci.*, **112**, 2670–2675, doi:10.1073/pnas.1409606112.
31 <http://www.pnas.org/content/112/9/2670.abstract>.
- 32 Moosa, C. S., and N. Tuana, 2014: Mapping a research agenda concerning gender and climate change: A
33 review of the literature. *Hypatia*, **29**, 677–694.
- 34 Morales-Hidalgo, D., S. N. Oswalt, and E. Somanathan, 2015: Status and trends in global primary forest,
35 protected areas, and areas designated for conservation of biodiversity from the Global Forest
36 Resources Assessment 2015. *For. Ecol. Manage.*, doi:10.1016/j.foreco.2015.06.011.
- 37 Moroni, S., 2018: Property as a human right and property as a special title. Rediscussing private
38 ownership of land. *Land use policy*, **70**, 273–280, doi:10.1016/J.LANDUSEPOL.2017.10.037.
39 <https://www.sciencedirect.com/science/article/pii/S0264837717303368> (Accessed April 11, 2019).
- 40 Mosnier, A., and Coauthors, 2014: Global food markets, trade and the cost of climate change adaptation.
41 *Food Secur.*, **6**, 29–44, doi:10.1007/s12571-013-0319-z. [http://link.springer.com/10.1007/s12571-](http://link.springer.com/10.1007/s12571-013-0319-z)
42 [013-0319-z](http://link.springer.com/10.1007/s12571-013-0319-z) (Accessed January 19, 2018).
- 43 Mottet, A., C. de Haan, A. Falcucci, G. Tempio, C. Opio, and P. Gerber, 2017: Livestock: On our plates
44 or eating at our table? A new analysis of the feed/food debate. *Glob. Food Sec.*, **14**, 1–8,

- 1 doi:10.1016/J.GFS.2017.01.001.
2 <https://www.sciencedirect.com/science/article/pii/S2211912416300013> (Accessed October 6, 2018).
- 3 Mouratiadou, I., and Coauthors, 2016: The impact of climate change mitigation on water demand for
4 energy and food: An integrated analysis based on the Shared Socioeconomic Pathways. *Environ.*
5 *Sci. Policy*, **64**, 48–58, doi:10.1016/J.ENVSCL.2016.06.007.
6 <https://www.sciencedirect.com/science/article/pii/S146290111630301X> (Accessed May 26, 2018).
- 7 Muller, A., and Coauthors, 2017: Strategies for feeding the world more sustainably with organic
8 agriculture. *Nat. Commun.*, **8**, doi:10.1038/s41467-017-01410-w.
- 9 Muller, C., J. Elliott, J. Chryssanthacopoulos, D. Deryng, C. Folberth, T. A. M. Pugh, and E. Schmid,
10 2015: Implications of climate mitigation for future agricultural production. *Environ. Res. Lett.*, **10**,
11 doi:12500410.1088/1748-9326/10/12/125004.
- 12 Murdiyarso, D., and Coauthors, 2015: The potential of Indonesian mangrove forests for global climate
13 change mitigation. *Nat. Clim. Chang.*, **5**, 1089–1092, doi:10.1038/NCLIMATE2734.
- 14 Myers, S. S., Zanobetti, A., Kloog, I., Huybers, P., Leakey, A. D., Bloom, A. J., 2014: Increasing CO 2
15 threatens human nutrition. *Nature*, **510**, 139. <https://www.nature.com/articles/nature13179>.
- 16 Myers, S. S., M. R. Smith, S. Guth, C. D. Golden, B. Vaitla, N. D. Mueller, A. D. Dangour, and P.
17 Huybers, 2017: Climate Change and Global Food Systems: Potential Impacts on Food Security and
18 Undernutrition. *Annu. Rev. Public Health*, **38**, 259–277, doi:10.1146/annurev-publhealth-031816-
19 044356. <http://www.annualreviews.org/doi/10.1146/annurev-publhealth-031816-044356> (Accessed
20 April 24, 2018).
- 21 Nachtergaele, F., 2008: *Mapping Land Use Systems at global and regional scales for Land Degradation*
22 *Assessment Analysis Version 1.0*.
- 23 Nakamura, A., and Coauthors, 2017: *Forests and Their Canopies: Achievements and Horizons in Canopy*
24 *Science*.
- 25 Nakicenovic, N., and R. Swart, 2000: *Special Report on Emissions Scenarios. Special Report on*
26 *Emissions Scenarios: Nature Publishing Group*. 612 pp.
- 27 Namubiru-Mwaura, E., 2014: *Land tenure and gender : approaches and challenges for strengthening*
28 *rural women's land rights*. World Bank, Washington, DC,
29 <https://openknowledge.worldbank.org/handle/10986/21033>.
- 30 Neill, B. C. O., 2004: Projections : An Application to Climate Change. 167–184.
- 31 Nelson, K. C., and B. H. J. de Jong, 2003: Making global initiatives local realities: carbon mitigation
32 projects in Chiapas, Mexico. *Glob. Environ. Chang.*, **13**, 19–30, doi:[https://doi.org/10.1016/S0959-](https://doi.org/10.1016/S0959-3780(02)00088-2)
33 [3780\(02\)00088-2](https://doi.org/10.1016/S0959-3780(02)00088-2). <http://www.sciencedirect.com/science/article/pii/S0959378002000882>.
- 34 Nepstad, D. C., W. Boyd, C. M. Stickler, T. Bezerra, and A. A. Azevedo, 2013: Responding to climate
35 change and the global land crisis: REDD+, market transformation and low-emissions rural
36 development. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, **368**, 20120167, doi:10.1098/rstb.2012.0167.
37 <http://www.ncbi.nlm.nih.gov/pubmed/23610173> (Accessed May 24, 2018).
- 38 Newbold, T., and Coauthors, 2015: Global effects of land use on local terrestrial biodiversity. *Nature*,
39 **520**, 45-, doi:10.1038/nature14324.
- 40 —, D. P. Tittensor, M. B. J. Harfoot, J. P. W. Scharlemann, and D. W. Purves, 2018: Non-linear
41 changes in modelled terrestrial ecosystems subjected to perturbations. *bioRxiv*, doi:10.1101/439059.
42 <https://www.biorxiv.org/content/biorxiv/early/2018/10/10/439059.full.pdf>.

- 1 Nicole, W., 2015: Pollinator Power: Nutrition Security Benefits of an Ecosystem Service. *Environ.*
2 *Health Perspect.*, **123**, A210–A215, doi:10.1289/ehp.123-A210.
3 <http://www.ncbi.nlm.nih.gov/pmc/articles/PMC4528997/>.
- 4 Nkhonjera, G. K., 2017: Understanding the impact of climate change on the dwindling water resources of
5 South Africa, focusing mainly on Olifants River basin: A review. *Environ. Sci. Policy*, **71**, 19–29,
6 doi:10.1016/J.ENVSCL.2017.02.004.
7 <https://www.sciencedirect.com/science/article/pii/S1462901116306608> (Accessed May 25, 2018).
- 8 Nolte, K., W. Chamberlain, and M. Giger, 2016: *International Land Deals for Agriculture Fresh insights*
9 *from the Land Matrix: Analytical Report II*. Bern, Montpellier, Hamburg, Pretoria, 1-56 pp.
10 [http://landmatrix.org/media/filer_public/ab/c8/abc8b563-9d74-4a47-9548-](http://landmatrix.org/media/filer_public/ab/c8/abc8b563-9d74-4a47-9548-cb59e4809b4e/land_matrix_2016_analytical_report_draft_ii.pdf)
11 [cb59e4809b4e/land_matrix_2016_analytical_report_draft_ii.pdf](http://landmatrix.org/media/filer_public/ab/c8/abc8b563-9d74-4a47-9548-cb59e4809b4e/land_matrix_2016_analytical_report_draft_ii.pdf).
- 12 van Noordwijk, M., and L. Brussaard, 2014: Minimizing the ecological footprint of food: closing yield
13 and efficiency gaps simultaneously? *Curr. Opin. Environ. Sustain.*, **8**, 62–70,
14 doi:10.1016/J.COSUST.2014.08.008.
15 <https://www.sciencedirect.com/science/article/pii/S1877343514000517> (Accessed May 24, 2018).
- 16 Noordwijk, M. Van, L. Tanika, and B. Lusiana, 2017: Flood risk reduction and flow buffering as
17 ecosystem services – Part 2 : Land use and rainfall intensity effects in Southeast Asia. *Hydrol. Earth*
18 *Syst. Sci.*, 2341–2360, doi:10.5194/hess-21-2341-2017.
- 19 Nordhaus, W., 2014: Estimates of the Social Cost of Carbon: Concepts and Results from the DICE-
20 2013R Model and Alternative Approaches. *J. Assoc. Environ. Resour. Econ.*, **1**, 273–312,
21 doi:10.1086/676035. <https://www.journals.uchicago.edu/doi/10.1086/676035> (Accessed April 3,
22 2019).
- 23 Norse, D., and X. Ju, 2015: Environmental costs of China’s food security. *Agric. Ecosyst. Environ.*, **209**,
24 5–14, doi:10.1016/J.AGEE.2015.02.014.
25 <https://www.sciencedirect.com/science/article/pii/S0167880915000699> (Accessed April 3, 2019).
- 26 Nowosad, J., T. F. Stepinski, and P. Netzel, 2018: Global assessment and mapping of changes in
27 mesoscale landscapes: 1992–2015. *Int. J. Appl. Earth Obs. Geoinf.*, doi:10.1016/j.jag.2018.09.013.
- 28 O’Neill, B. C., E. Kriegler, K. Riahi, K. L. Ebi, S. Hallegatte, T. R. Carter, R. Mathur, and D. P. van
29 Vuuren, 2014: A new scenario framework for climate change research: the concept of shared
30 socioeconomic pathways. *Clim. Change*, **122**, 387–400, doi:10.1007/s10584-013-0905-2.
- 31 OECD, 2014: *Social Institutions and Gender Index (SIGI)*. Paris, [https://www.oecd.org/dev/development-](https://www.oecd.org/dev/development-gender/BrochureSIGI2015-web.pdf)
32 [gender/BrochureSIGI2015-web.pdf](https://www.oecd.org/dev/development-gender/BrochureSIGI2015-web.pdf).
- 33 Ogle, S. M., and Coauthors, 2018: Delineating managed land for reporting national greenhouse gas
34 emissions and removals to the United Nations framework convention on climate change. *Carbon*
35 *Balance Manag.*, **13**, doi:10.1186/s13021-018-0095-3. <https://doi.org/10.1186/s13021-018-0095-3>.
- 36 Ordway, E. M., G. P. Asner, and E. F. Lambin, 2017: Deforestation risk due to commodity crop
37 expansion in sub-Saharan Africa. *Environ. Res. Lett.*, **12**, 044015, doi:10.1088/1748-9326/aa6509.
38 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=4/a=044015?key=crossref.bbd762c2bf0b22e181e42a322b781549)
39 [9326/12/i=4/a=044015?key=crossref.bbd762c2bf0b22e181e42a322b781549](http://stacks.iop.org/1748-9326/12/i=4/a=044015?key=crossref.bbd762c2bf0b22e181e42a322b781549) (Accessed March 21,
40 2019).
- 41 Osborne, T. M., and T. R. Wheeler, 2013: Evidence for a climate signal in trends of global crop yield
42 variability over the past 50 years. *Environ. Res. Lett.*, **8**, 024001, doi:10.1088/1748-
43 [9326/8/2/024001](http://stacks.iop.org/1748-9326/8/2/024001). [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/8/2/024001)
44 [9326/8/2/024001](http://stacks.iop.org/1748-9326/8/2/024001)?key=crossref.17c24fb9d5a7de2b6c71cc62ace11600 (Accessed November 2,
45 2018).

- 1 Ostrom, E., and M. Cox, 2010: Moving beyond panaceas: a multi-tiered diagnostic approach for social-
2 ecological analysis. *Environ. Conserv.*, **37**, 451–463, doi:DOI: 10.1017/S0376892910000834.
3 [https://www.cambridge.org/core/article/moving-beyond-panaceas-a-multitiered-diagnostic-
4 approach-for-socialecological-analysis/F4870A21ED502BB7D9A1784CF2B9E100](https://www.cambridge.org/core/article/moving-beyond-panaceas-a-multitiered-diagnostic-approach-for-socialecological-analysis/F4870A21ED502BB7D9A1784CF2B9E100).
- 5 Padmanaba, M., and R. T. Corlett, 2014: Minimizing risks of invasive alien plant species in tropical
6 production forest management. *Forests*, **5**, 1982–1998, doi:10.3390/f5081982.
- 7 Paillet, Y., and Coauthors, 2010: *Biodiversity Differences between Managed and Unmanaged Forests:
8 Meta-Analysis of Species Richness in Europe*. 101–112 pp.
- 9 Di Paola, A., L. Caporaso, F. Di Paola, A. Bombelli, I. Vasenev, O. V. Nesterova, S. Castaldi, and R.
10 Valentini, 2018: The expansion of wheat thermal suitability of Russia in response to climate change.
11 *Land use policy*, **78**, 70–77, doi:10.1016/J.LANDUSEPOL.2018.06.035.
12 <https://www.sciencedirect.com/science/article/pii/S0264837717315880> (Accessed March 21, 2019).
- 13 Parfitt, J., M. Barthel, and S. Macnaughton, 2010: Food waste within food supply chains: quantification
14 and potential for change to 2050. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, **365**, 3065–3081,
15 doi:10.1098/rstb.2010.0126. <https://www.ncbi.nlm.nih.gov/pubmed/20713403>.
- 16 Parker, W. S., 2013: Ensemble modeling, uncertainty and robust predictions. *Wiley Interdiscip. Rev.
17 Chang.*, **4**, 213–223, doi:10.1002/wcc.220.
- 18 Parr, C. L., C. E. R. R. Lehmann, W. J. Bond, W. A. Hoffmann, and A. N. Andersen, 2014: Tropical
19 grassy biomes: misunderstood, neglected, and under threat. *Trends Ecol. Evol.*, **29**, 205–213,
20 doi:10.1016/j.tree.2014.02.004. <http://www.ncbi.nlm.nih.gov/pubmed/24629721> (Accessed
21 September 19, 2018).
- 22 Pawson, S. M., A. Brin, E. G. Brockerhoff, D. Lamb, T. W. Payn, A. Paquette, and J. A. Parrotta, 2013:
23 Plantation forests, climate change and biodiversity. *Biodivers. Conserv.*, **22**, 1203–1227,
24 doi:10.1007/s10531-013-0458-8. <http://link.springer.com/10.1007/s10531-013-0458-8> (Accessed
25 September 20, 2018).
- 26 Payn, T., and Coauthors, 2015: Changes in planted forests and future global implications. *For. Ecol.
27 Manage.*, **352**, 57–67, doi:10.1016/J.FORECO.2015.06.021.
28 <https://www.sciencedirect.com/science/article/pii/S0378112715003473> (Accessed September 25,
29 2018).
- 30 Pedrozo-Acuña, A., R. Damania, M. A. Laverde-Barajas, and D. Mira-Salama, 2015: Assessing the
31 consequences of sea-level rise in the coastal zone of Quintana Roo, México: the costs of inaction. *J.
32 Coast. Conserv.*, **19**, 227–240, doi:10.1007/s11852-015-0383-y.
33 <http://link.springer.com/10.1007/s11852-015-0383-y> (Accessed April 3, 2019).
- 34 Pereira, H. M., and Coauthors, 2010: Scenarios for Global Biodiversity in the 21st Century. *Science (80-
35).*, **330**, 1496–1501, doi:10.1126/science.1196624.
- 36 Perugini, L., and Coauthors, 2017: Biophysical effects on temperature and precipitation due to land cover
37 change Biophysical effects on temperature and precipitation due to land cover change. *Environ. Res.
38 Lett.*, **12**, 1–21, doi:<https://doi.org/10.1088/1748-9326/aa6b3f>.
- 39 Peterson, E. E., S. A. Cunningham, M. Thomas, S. Collings, G. D. Bonnett, and B. Harch, 2017: An
40 assessment framework for measuring agroecosystem health. *Ecol. Indic.*, **79**, 265–275,
41 doi:10.1016/j.ecolind.2017.04.002. [https://ac.els-cdn.com/S1470160X17301772/1-s2.0-
42 S1470160X17301772-main.pdf?_tid=f3112b54-efed-11e7-a298-
43 00000aab0f02&acdnat=1514919272_41be4aa77bfb5817782eb53d77020635](https://ac.els-cdn.com/S1470160X17301772/1-s2.0-S1470160X17301772-main.pdf?_tid=f3112b54-efed-11e7-a298-00000aab0f02&acdnat=1514919272_41be4aa77bfb5817782eb53d77020635) (Accessed January 2,
44 2018).

- 1 Pham, P., P. Doneys, and D. L. Doane, 2016: Changing livelihoods, gender roles and gender hierarchies:
2 The impact of climate, regulatory and socio-economic changes on women and men in a Co Tu
3 community in Vietnam. *Womens. Stud. Int. Forum*, **54**, 48–56, doi:10.1016/J.WSIF.2015.10.001.
4 <https://www.sciencedirect.com/science/article/pii/S0277539515001569> (Accessed April 3, 2019).
- 5 Pimm, S. L., and Coauthors, 2014: The biodiversity of species and their rates of extinction, distribution,
6 and protection. *Science (80-.)*, **344**, 1246752–1246752, doi:10.1126/science.1246752.
7 <http://www.sciencemag.org/cgi/doi/10.1126/science.1246752>.
- 8 Pingoud, K., T. Ekholm, R. Sievänen, S. Huuskonen, and J. Hynynen, 2018: Trade-offs between forest
9 carbon stocks and harvests in a steady state – A multi-criteria analysis. *J. Environ. Manage.*, **210**,
10 96–103, doi:10.1016/J.JENVMAN.2017.12.076.
11 <https://www.sciencedirect.com/science/article/pii/S0301479717312641> (Accessed April 25, 2018).
- 12 Pizer, W., and Coauthors, 2014: Using and improving the social cost of carbon. *Science (80-.)*, **346**,
13 1189–1190, doi:10.1126/science.1259774. <http://science.sciencemag.org/content/346/6214/1189>.
- 14 Poeplau, C., and A. Don, 2015: Carbon sequestration in agricultural soils via cultivation of cover crops –
15 A meta-analysis. *Agric. Ecosyst. Environ.*, **200**, 33–41, doi:10.1016/J.AGEE.2014.10.024.
16 <https://www.sciencedirect.com/science/article/pii/S0167880914004873> (Accessed January 19,
17 2018).
- 18 —, —, L. Vesterdal, J. Leifeld, B. Van Wesemael, J. Schumacher, and A. Gensior, 2011: Temporal
19 dynamics of soil organic carbon after land-use change in the temperate zone - carbon response
20 functions as a model approach. *Glob. Chang. Biol.*, **17**, 2415–2427, doi:10.1111/j.1365-
21 2486.2011.02408.x. <http://doi.wiley.com/10.1111/j.1365-2486.2011.02408.x> (Accessed September
22 24, 2018).
- 23 Poore, J., and T. Nemecek, 2018: Reducing food’s environmental impacts through producers and
24 consumers. *Science (80-.)*, **360**, 987–992, doi:10.1126/science.aag0216.
25 <http://science.sciencemag.org/content/sci/360/6392/987.full.pdf>.
- 26 Popp, A., and Coauthors, 2014: Land-use protection for climate change mitigation. *Nat. Clim. Chang.*, **4**,
27 1095–1098, doi:10.1038/nclimate2444.
- 28 —, and Coauthors, 2016: Land-use futures in the shared socio-economic pathways. *Glob. Environ.*
29 *Chang.*, **42**, doi:<http://dx.doi.org/10.1016/j.gloenvcha.2016.10.002>.
30 <https://www.sciencedirect.com/science/article/pii/S0959378016303399> (Accessed May 26, 2018).
- 31 Porter, J. R., L. Xie, A. J. Challinor, K. Cochrane, S. M. Howden, M. M. Iqbal, D. B. Lobell, and M. I.
32 Travasso, 2014: Food security and food production systems. *Climate Change 2014: Impacts,*
33 *Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group*
34 *II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, Cambridge
35 University Press, 485–533.
- 36 Porter, S. D., D. S. Reay, P. Higgins, and E. Bomberg, 2016: A half-century of production-phase
37 greenhouse gas emissions from food loss & waste in the global food supply chain. *Sci. Total*
38 *Environ.*, **571**, 721–729, doi:10.1016/J.SCITOTENV.2016.07.041.
39 <https://www.sciencedirect.com/science/article/pii/S0048969716314863> (Accessed May 24, 2018).
- 40 Potapov, P., and Coauthors, 2017: The last frontiers of wilderness: {Tracking} loss of intact forest
41 landscapes from 2000 to 2013. *Sci. Adv.*, **3**, e1600821, doi:10.1126/sciadv.1600821.
42 <http://advances.sciencemag.org/content/3/1/e1600821> (Accessed March 24, 2017).
- 43 Pradhan, P., M. K. B. Lüdeke, D. E. Reusser, and J. P. Kropp, 2013: Embodied crop calories in animal
44 products. *Environ. Res. Lett.*, **8**, doi:10.1088/1748-9326/8/4/044044.

- 1 ———, D. E. Reusser, and J. P. Kropp, 2014: Food Self-Sufficiency across Scales: How Local Can We
2 Go? **15**, 9779.
- 3 Pravalie, R., 2016: Drylands extent and environmental issues. A global approach. *Earth-Science Rev.*,
4 **161**, 259–278, doi:10.1016/j.earscirev.2016.08.003.
- 5 Prestele, R., and Coauthors, 2016: Hotspots of uncertainty in land-use and land-cover change projections:
6 a global-scale model comparison. *Glob. Chang. Biol.*, **22**, 3967–3983, doi:10.1111/gcb.13337.
- 7 Pugh, T. A. M., C. Mueller, J. Elliott, D. Deryng, C. Folberth, S. Olin, E. Schmid, and A. Arneeth, 2016:
8 Climate analogues suggest limited potential for intensification of production on current croplands
9 under climate change. *Nat. Commun.*, **7**, doi:1260810.1038/ncomms12608.
- 10 Pugh, T. A. M., M. Lindeskog, B. Smith, B. Poulter, A. Arneeth, V. Haverd, and L. Calle, 2019: Role of
11 forest regrowth in global carbon sink dynamics. *Proc. Natl. Acad. Sci.*, 201810512,
12 doi:10.1073/pnas.1810512116.
13 <https://www.pnas.org/content/pnas/early/2019/02/12/1810512116.full.pdf>.
- 14 Putz, F. E., and K. H. Redford, 2010: The importance of defining “Forest”: Tropical forest degradation,
15 deforestation, long-term phase shifts, and further transitions. *Biotropica*, doi:10.1111/j.1744-
16 7429.2009.00567.x.
- 17 Le Quere, C., and Coauthors, 2015: Global Carbon Budget 2015. *Earth Syst. Sci. Data*, **7**, 349–396,
18 doi:10.5194/essd-7-349-2015.
- 19 ———, and Coauthors, 2018: Global Carbon Budget 2017. *Earth Syst. Sci. Data*, **10**, 405–448,
20 doi:10.5194/essd-10-405-2018.
- 21 Le Quéré, C., and Coauthors, 2013: The global carbon budget 1959–2011. *Earth Syst. Sci. Data*, **5**, 165–
22 185, doi:10.5194/essd-5-165-2013. <http://www.earth-syst-sci-data.net/5/165/2013/> (Accessed April
23 24, 2018).
- 24 Le Quéré, C., and Coauthors, 2018: Global Carbon Budget 2018. *Earth Syst. Sci. Data Discuss.*, 1–3,
25 doi:10.5194/essd-2018-120. <https://www.earth-syst-sci-data-discuss.net/essd-2018-120/> (Accessed
26 November 1, 2018).
- 27 Raffensperger, C., and J. A. Tickner, 1999: Introduction: to Foresee and Forestall. *Protecting public
28 health & the environment: implementing the precautionary principle CN - GE105 .P76 1999*, Island
29 Press, Washington, D.C, 1–11.
- 30 Raiten, D. J., and A. M. Aimone, 2017: The intersection of climate/environment, food, nutrition and
31 health: crisis and opportunity. *Curr. Opin. Biotechnol.*, **44**, 52–62,
32 doi:10.1016/J.COPBIO.2016.10.006.
33 <https://www.sciencedirect.com/science/article/pii/S0958166916302336> (Accessed May 25, 2018).
- 34 Ramankutty, N., A. T. Evan, C. Monfreda, and J. A. Foley, 2008: Farming the planet: 1. {Geographic}
35 distribution of global agricultural lands in the year 2000. *Global Biogeochem. Cycles*, **22**, GB1003,
36 doi:10.1029/2007GB002952.
- 37 ———, Z. Mehrabi, K. Waha, L. Jarvis, C. Kremen, M. Herrero, and L. H. Rieseberg, 2018: Trends in
38 Global Agricultural Land Use: Implications for Environmental Health and Food Security. *Annu.
39 Rev. Plant Biol.*, **69**, 789–815, doi:10.1146/annurev-arplant-042817-040256.
40 <http://www.annualreviews.org/doi/10.1146/annurev-arplant-042817-040256> (Accessed March 21,
41 2019).
- 42 Randerson, J. T., and Coauthors, 2009: Systematic assessment of terrestrial biogeochemistry in coupled
43 climate-carbon models. *Glob. Chang. Biol.*, **15**, 2462–2484, doi:10.1111/j.1365-2486.2009.01912.x.

- 1 Rao, N., 2017: Assets, Agency and Legitimacy: Towards a Relational Understanding of Gender Equality
2 Policy and Practice. *World Dev.*, **95**, 43–54, doi:10.1016/J.WORLDDEV.2017.02.018.
3 <https://www.sciencedirect.com/science/article/abs/pii/S0305750X15308810> (Accessed April 18,
4 2019).
- 5 Rao, Y., M. Zhou, G. Ou, D. Dai, L. Zhang, Z. Zhang, X. Nie, and C. Yang, 2018: Integrating ecosystem
6 services value for sustainable land-use management in semi-arid region. *J. Clean. Prod.*, **186**, 662–
7 672, doi:10.1016/J.JCLEPRO.2018.03.119.
8 <https://www.sciencedirect.com/science/article/pii/S0959652618307807> (Accessed April 25, 2018).
- 9 Ravi, S., D. D. Breshears, T. E. Huxman, and P. D’Odorico, 2010: Land degradation in drylands:
10 Interactions among hydrologic–aeolian erosion and vegetation dynamics. *Geomorphology*, **116**,
11 236–245, doi:10.1016/j.geomorph.2009.11.023.
12 <http://linkinghub.elsevier.com/retrieve/pii/S0169555X09005108> (Accessed April 21, 2018).
- 13 Ravnborg, H. M., R. Spichiger, R. B. Broegaard, and R. H. Pedersen, 2016: Land Governance, Gender
14 Equality and Development: Past Achievements and Remaining Challenges. *J. Int. Dev.*, **28**, 412–
15 427, doi:10.1002/jid.3215. <https://doi.org/10.1002/jid.3215>.
- 16 Ray, D. K., N. Ramankutty, N. D. Mueller, P. C. West, and J. A. Foley, 2012: Recent patterns of crop
17 yield growth and stagnation. *Nat. Commun.*, **3**, doi:10.1038/ncomms2296.
- 18 Reed, M., and L. C. Stringer, 2015: *Climate change and desertification: Anticipating, assessing &*
19 *adapting to future change in drylands*. Agropolis International.
- 20 Resurrección, B. P., 2013: Persistent women and environment linkages in climate change and sustainable
21 development agendas. *Womens. Stud. Int. Forum*, **40**, 33–43, doi:10.1016/J.WSIF.2013.03.011.
22 <https://www.sciencedirect.com/science/article/pii/S0277539513000599?via%3Dihub> (Accessed
23 April 3, 2019).
- 24 Reyer, C., M. Guericke, and P. L. Ibsch, 2009: Climate change mitigation via afforestation, reforestation
25 and deforestation avoidance: and what about adaptation to environmental change? *New For.*, **38**,
26 15–34, doi:10.1007/s11056-008-9129-0. <https://doi.org/10.1007/s11056-008-9129-0>.
- 27 Riahi, K., and Coauthors, 2015: Locked into Copenhagen pledges — Implications of short-term emission
28 targets for the cost and feasibility of long-term climate goals. *Technol. Forecast. Soc. Change*, **90**,
29 8–23, doi:<https://doi.org/10.1016/j.techfore.2013.09.016>.
30 <http://www.sciencedirect.com/science/article/pii/S0040162513002539>.
- 31 Riahi, K., and Coauthors, 2017: The Shared Socioeconomic Pathways and their energy, land use, and
32 greenhouse gas emissions implications: An overview. *Glob. Environ. Chang.*, **42**, 153–168.
- 33 Richards, D. R., and D. A. Friess, 2016: Rates and drivers of mangrove deforestation in Southeast Asia,
34 2000–2012. *Proc. Natl. Acad. Sci. U. S. A.*, **113**, 344–349, doi:10.1073/pnas.1510272113.
35 <http://www.ncbi.nlm.nih.gov/pubmed/26712025> (Accessed March 21, 2019).
- 36 Ricke, K., L. Drouet, K. Caldeira, and M. Tavoni, 2018: Country-level social cost of carbon. *Nat. Clim.*
37 *Chang.*, **8**, 895–900, doi:10.1038/s41558-018-0282-y. <https://doi.org/10.1038/s41558-018-0282-y>.
- 38 Ringler, C., and R. Lawford, 2013: The nexus across water, energy, land and food (WELF): potential for
39 improved resource use efficiency? *Curr. Opin. Environ. Sustain.*, **5**, 617–624,
40 doi:10.1016/J.COSUST.2013.11.002.
41 <https://www.sciencedirect.com/science/article/pii/S1877343513001504> (Accessed May 24, 2018).
- 42 Robinson, D. A., and Coauthors, 2017: Modelling feedbacks between human and natural processes in the
43 land system. *Earth Syst. Dyn. Discuss.*, doi:<https://doi.org/10.5194/esd-2017-68>.

- 1 Rodd Myers, Anna JP Sanders, Anne M Larson, Rut Dini Prasti H, Ashwin Ravikumar, 2016:
2 *Analyzing multilevel governance in Indonesia: lessons for REDD+ from the study of landuse change*
3 *in Central and West Kalimantan*. ix + 69 pp. [https://xs.glgoo.net/scholar?hl=zh-](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=Analyzing+multilevel+governance+in+Indonesia%3A+lessons+for+REDD%2B+from+the+study+of+landuse+change+in+Central+and+West+Kalimantan.&btnG=)
4 [CN&as_sdt=0%2C5&q=Analyzing+multilevel+governance+in+Indonesia%3A+lessons+for+REDD](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=Analyzing+multilevel+governance+in+Indonesia%3A+lessons+for+REDD%2B+from+the+study+of+landuse+change+in+Central+and+West+Kalimantan.&btnG=)
5 [%2B+from+the+study+of+landuse+change+in+Central+and+West+Kalimantan.&btnG=.](https://xs.glgoo.net/scholar?hl=zh-CN&as_sdt=0%2C5&q=Analyzing+multilevel+governance+in+Indonesia%3A+lessons+for+REDD%2B+from+the+study+of+landuse+change+in+Central+and+West+Kalimantan.&btnG=)
- 6 Rodriguez-Labajos, B., 2013: Climate change, ecosystem services, and costs of action and inaction:
7 scoping the interface. *Wiley Interdiscip. Rev. Chang.*, **4**, 555–573, doi:10.1002/wcc.247.
- 8 Rogelj, J., and Coauthors, 2018a: Mitigation pathways compatible with 1.5°C in the context of
9 sustainable development. *Global Warming of 1.5 °C an IPCC special report on the impacts of*
10 *global warming of 1.5 °C above pre-industrial levels and related global greenhouse gas emission*
11 *pathways, in the context of strengthening the global response to the threat of climate change*, V.
12 Masson-Delmotte et al., Eds., World Meteorological Organization, Geneva, Switzerland
13 <http://www.ipcc.ch/report/sr15/>.
- 14 Rogelj, J., and Coauthors, 2018b: Scenarios towards limiting global mean temperature increase below 1.5
15 degrees C. *Nat. Clim. Chang.*, **8**, 325–+, doi:10.1038/s41558-018-0091-3.
- 16 Rook, G. A., 2013: Regulation of the immune system by biodiversity from the natural environment: An
17 ecosystem service essential to health. *Proc. Natl. Acad. Sci.*, **110**, 18360–18367,
18 doi:10.1073/pnas.1313731110. <https://www.pnas.org/content/pnas/110/46/18360.full.pdf>.
- 19 Rööös, E., B. Bajželj, P. Smith, M. Patel, D. Little, and T. Garnett, 2017: Greedy or needy? Land use and
20 climate impacts of food in 2050 under different livestock futures. *Glob. Environ. Chang.*, **47**, 1–12,
21 doi:10.1016/J.GLOENVCHA.2017.09.001.
22 <https://www.sciencedirect.com/science/article/pii/S0959378016306872> (Accessed May 25, 2018).
- 23 Rosa, I. M. D. I. M. D., and Coauthors, 2017: Multiscale scenarios for nature futures. *Nat. Ecol. Evol.*, **1**,
24 1416–1419, doi:10.1038/s41559-017-0273-9. <http://www.nature.com/articles/s41559-017-0273-9>.
- 25 Rose, S. K., 2014: Integrated assessment modeling of climate change adaptation in forestry and pasture
26 land use: A review. *Energy Econ.*, **46**, 548–554, doi:10.1016/J.ENECO.2014.09.018.
27 <https://www.sciencedirect.com/science/article/pii/S014098831400228X> (Accessed April 12, 2019).
- 28 Rosen, R. A., and E. Guenther, 2015: The economics of mitigating climate change: What can we know?
29 *Technol. Forecast. Soc. Change*, **91**, 93–106, doi:10.1016/J.TECHFORE.2014.01.013.
30 <https://www.sciencedirect.com/science/article/pii/S0040162514000468> (Accessed November 2,
31 2018).
- 32 Rosenzweig, C., and P. Neofotis, 2013: Detection and attribution of anthropogenic climate change
33 impacts. *Wiley Interdiscip. Rev. Chang.*, **4**, 121–150, doi:10.1002/wcc.209.
- 34 Rosenzweig, C., and Coauthors, 2014: Assessing agricultural risks of climate change in the 21st century
35 in a global gridded crop model intercomparison. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 3268–3273,
36 doi:10.1073/pnas.1222463110.
- 37 Rounsevell, M. D. A., and M. J. Metzger, 2010: Developing qualitative scenario storylines for
38 environmental change assessment. *Wiley Interdiscip. Rev. Clim. Chang.*, **1**, 606–619,
39 doi:10.1002/wcc.63.
- 40 Rounsevell, M. D. A., and Coauthors, 2006: A coherent set of future land use change scenarios for
41 Europe. *Agric. Ecosyst. Environ.*, **114**, 57–68, doi:10.1016/j.agee.2005.11.027.
- 42 Rounsevell, M. D. A., and Coauthors, 2014: Towards decision-based global land use models for
43 improved understanding of the Earth system. *Earth Syst. Dyn.*, **5**, 117–137, doi:10.5194/esd-5-117-
44 2014. <http://www.earth-syst-dynam.net/5/117/2014/>.

- 1 Roy, J., and Coauthors, 2018: Sustainable Development , Poverty Eradication and Reducing Inequalities.
2 *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of 1.5 °C*
3 *above pre-industrial levels and related global greenhouse gas emission pathways, in the context of*
4 *strengthening the global response to the threat of climate change*, V. Masson-Delmotte et al., Eds.,
5 World Meteorological Organization, Geneva, Switzerland <http://www.ipcc.ch/report/sr15/>.
- 6 Rulli, M. C., A. Savioli, and P. D’Odorico, 2012: Global land and water grabbing. *Pnas*, **110**, 892–897,
7 doi:10.1073/pnas.1213163110/-DCSupplemental.www.pnas.org/cgi/doi/10.1073/pnas.1213163110.
- 8 Runting, R. K., and Coauthors, 2017: Incorporating climate change into ecosystem service assessments
9 and decisions: a review. *Glob. Chang. Biol.*, **23**, 28–41, doi:10.1111/gcb.13457.
10 <http://dx.doi.org/10.1111/gcb.13457>.
- 11 Ryan, C. M., R. Pritchard, L. McNicol, M. Owen, J. A. Fisher, and C. Lehmann, 2016: Ecosystem
12 services from southern African woodlands and their future under global change. *Philos. Trans. R.*
13 *Soc. B-Biological Sci.*, **371**, doi:2015031210.1098/rstb.2015.0312.
- 14 Salmon, G., N. Teufel, I. Baltenweck, M. van Wijk, L. Claessens, and K. Marshall, 2018: Trade-offs in
15 livestock development at farm level: Different actors with different objectives. *Glob. Food Sec.*,
16 doi:10.1016/J.GFS.2018.04.002.
17 <https://www.sciencedirect.com/science/article/pii/S2211912417301372> (Accessed May 25, 2018).
- 18 Salvati, L., and M. Carlucci, 2014: Zero Net Land Degradation in Italy: The role of socioeconomic and
19 agro-forest factors. *J. Environ. Manage.*, **145**, 299–306, doi:10.1016/J.JENVMAN.2014.07.006.
20 <https://www.sciencedirect.com/science/article/pii/S0301479714003387> (Accessed April 23, 2018).
- 21 Salvati, L., A. Sabbi, D. Smiraglia, and M. Zitti, 2014: Does forest expansion mitigate the risk of
22 desertification? Exploring soil degradation and land-use changes in a Mediterranean country. *Int.*
23 *For. Rev.*, **16**, 485–496, doi:10.1505/146554814813484149.
- 24 Santangeli, A., T. Toivonen, F. M. Pouzols, M. Pogson, A. Hastings, P. Smith, and A. Moilanen, 2016:
25 Global change synergies and trade-offs between renewable energy and biodiversity. *Glob. Chang.*
26 *Biol. Bioenergy*, **8**, doi:10.1111/gcbb.12299.
- 27 Santilli, G., C. Vendittozzi, C. Cappelletti, S. Battistini, and P. Gessini, 2018: CubeSat constellations for
28 disaster management in remote areas. *Acta Astronaut.*, **145**, 11–17,
29 doi:10.1016/j.actaastro.2017.12.050. <https://doi.org/10.1016/j.actaastro.2017.12.050>.
- 30 Sanz-Sanchez, M.-J., and Coauthors, 2017: *Sustainable Land Management contribution to successful*
31 *land-based climate change adaptation and mitigation*.
- 32 Schaeffer, M., L. Gohar, E. Kriegler, J. Lowe, K. Riahi, and D. van Vuuren, 2015: Mid- and long-term
33 climate projections for fragmented and delayed-action scenarios. *Technol. Forecast. Soc. Change*,
34 **90**, 257–268, doi:<https://doi.org/10.1016/j.techfore.2013.09.013>.
35 <http://www.sciencedirect.com/science/article/pii/S0040162513002424>.
- 36 Schanes, K., K. Dobernig, and B. Gözet, 2018: Food waste matters - A systematic review of household
37 food waste practices and their policy implications. *J. Clean. Prod.*, **182**, 978–991,
38 doi:10.1016/J.JCLEPRO.2018.02.030.
39 <https://www.sciencedirect.com/science/article/pii/S0959652618303366> (Accessed April 3, 2019).
- 40 Schauburger, B., and Coauthors, 2017: Consistent negative response of US crops to high temperatures in
41 observations and crop models. *Nat. Commun.*, **8**, doi:10.1038/ncomms13931.
- 42 Scheidel, A., and C. Work, 2018: Forest plantations and climate change discourses: New powers of
43 ‘green’ grabbing in Cambodia. *Land use policy*, **77**, 9–18, doi:10.1016/j.landusepol.2018.04.057.
44 <https://doi.org/10.1016/j.landusepol.2018.04.057>.

- 1 Schepaschenko, D., and Coauthors, 2015: Development of a global hybrid forest mask through the
2 synergy of remote sensing, crowdsourcing and FAO statistics. *Remote Sens. Environ.*, **162**, 208–
3 220. <https://www.sciencedirect.com/science/article/pii/S0034425715000644> (Accessed May 26,
4 2018).
- 5 Schipper, L. A., R. L. Parfitt, S. Fraser, R. A. Littler, W. T. Baisden, and C. Ross, 2014: Soil order and
6 grazing management effects on changes in soil C and N in New Zealand pastures. *Agric. Ecosyst.*
7 *Environ.*, **184**, 67–75, doi:10.1016/J.AGEE.2013.11.012.
8 <https://www.sciencedirect.com/science/article/pii/S0167880913004064> (Accessed April 3, 2019).
- 9 Schlenker, W., and D. B. Lobell, 2010: Robust negative impacts of climate change on African agriculture.
10 *Environ. Res. Lett.*, **5**, 14010. <http://stacks.iop.org/1748-9326/5/i=1/a=014010>.
- 11 Schlesinger, W. H., 2018: Are wood pellets a green fuel? *Science*, **359**, 1328–1329,
12 doi:10.1126/science.aat2305. <http://www.ncbi.nlm.nih.gov/pubmed/29567691> (Accessed April 25,
13 2018).
- 14 Schmidt, C. G., K. Foerstl, and B. Schaltenbrand, 2017: The Supply Chain Position Paradox: Green
15 Practices and Firm Performance. *J. Supply Chain Manag.*, **53**, 3–25, doi:10.1111/jscm.12113.
16 <http://doi.wiley.com/10.1111/jscm.12113> (Accessed April 18, 2019).
- 17 Schneider, F., and T. Buser, 2018: Promising degrees of stakeholder interaction in research for
18 sustainable development. *Sustain. Sci.*, **13**, 129–142, doi:10.1007/s11625-017-0507-4.
19 <http://link.springer.com/10.1007/s11625-017-0507-4> (Accessed April 18, 2019).
- 20 Scholes, R., and Coauthors, 2018: *IPBES: Summary for policymakers of the thematic assessment report*
21 *on land degradation and restoration of the Intergovernmental Science-Policy Platform on*
22 *Biodiversity and Ecosystem Services*. Bonn, Germany, [] pages pp.
23 [https://naturesciences.ch/uuid/8c9635a6-9b7b-50bc-b615-](https://naturesciences.ch/uuid/8c9635a6-9b7b-50bc-b615-38adddb915cc?r=20180524162801_1527108923_feac0ee0-02f9-589f-b41f-7b3dcba029c3)
24 [38adddb915cc?r=20180524162801_1527108923_feac0ee0-02f9-589f-b41f-7b3dcba029c3](https://naturesciences.ch/uuid/8c9635a6-9b7b-50bc-b615-38adddb915cc?r=20180524162801_1527108923_feac0ee0-02f9-589f-b41f-7b3dcba029c3).
- 25 Schröter, M., and Coauthors, 2018: Interregional flows of ecosystem services: Concepts, typology and
26 four cases. *Ecosyst. Serv.*, doi:10.1016/j.ecoser.2018.02.003.
- 27 Schulte, R. P. O., R. E. Creamer, T. Donnellan, N. Farrelly, R. Fealy, C. O’Donoghue, and D.
28 O’hUallachain, 2014: Functional land management: A framework for managing soil-based
29 ecosystem services for the sustainable intensification of agriculture. *Environ. Sci. Policy*, **38**, 45–58,
30 doi:10.1016/J.ENVSCI.2013.10.002.
31 <https://www.sciencedirect.com/science/article/pii/S1462901113002104> (Accessed January 19,
32 2018).
- 33 Schut, M., and Coauthors, 2016: Sustainable intensification of agricultural systems in the Central African
34 Highlands: The need for institutional innovation. *Agric. Syst.*, **145**, 165–176,
35 doi:10.1016/J.AGSY.2016.03.005.
36 <https://www.sciencedirect.com/science/article/pii/S0308521X16300440> (Accessed May 24, 2018).
- 37 Schweikert, A., P. Chinowsky, X. Espinet, and M. Tarbert, 2014: Climate Change and Infrastructure
38 Impacts: Comparing the Impact on Roads in ten Countries through 2100. *Procedia Eng.*, **78**, 306–
39 316, doi:<https://doi.org/10.1016/j.proeng.2014.07.072>.
40 <http://www.sciencedirect.com/science/article/pii/S1877705814010595>.
- 41 Searchinger, T. D., and Coauthors, 2015: High carbon and biodiversity costs from converting Africa’s
42 wet savannahs to cropland. *Nat. Clim. Chang.*, **5**, 481–486, doi:10.1038/nclimate2584.
- 43 ———, T. Beringer, and A. Strong, 2017: Does the world have low-carbon bioenergy potential from the
44 dedicated use of land? *Energy Policy*, **110**, 434–446, doi:10.1016/j.enpol.2017.08.016.

- 1 Searle, S., and C. Malins, 2015: A reassessment of global bioenergy potential in 2050. *GCB Bioenergy*, **7**,
2 328–336, doi:10.1111/gcbb.12141. <http://doi.wiley.com/10.1111/gcbb.12141> (Accessed April 25,
3 2018).
- 4 Seidl, R., and Coauthors, 2017: Forest disturbances under climate change. *Nat. Clim. Chang.*,
5 doi:10.1038/nclimate3303.
- 6 Seneviratne, S. I., and Coauthors, 2018: Climate extremes, land-climate feedbacks and land-use forcing at
7 1.5 degrees C. *Philos. Trans. R. Soc. a-Mathematical Phys. Eng. Sci.*, **376**,
8 doi:2016045010.1098/rsta.2016.0450.
- 9 Seto, K. C., and A. Reenberg, 2014: *Rethinking global land use in an urban era*. The MIT Press,
10 <https://mitpress.mit.edu/books/rethinking-global-land-use-urban-era#.WwezDqvuvY4>.mendeley
11 (Accessed May 25, 2018).
- 12 —, and N. Ramankutty, 2016: Hidden linkages between urbanization and food systems. *Science* (80-
13), **352**, 943–945, doi:10.1126/science.aaf7439. <http://science.sciencemag.org/content/352/6288/943>
14 (Accessed May 30, 2016).
- 15 Seto, K. C., B. Guneralp, and L. R. Hutya, 2012: Global forecasts of urban expansion to 2030 and direct
16 impacts on biodiversity and carbon pools. *Proc. Natl. Acad. Sci.*, **109**, 16083–16088,
17 doi:10.1073/pnas.1211658109. <http://www.pnas.org/content/early/2012/09/11/1211658109>
18 (Accessed September 24, 2012).
- 19 Settele, J., R. Scholes, R. Betts, S. Bunn, P. Leadley, D. Nepstad, J. T. Overpeck, and M. A. Taboad,
20 2014: Terrestrial and inland water systems. *Climate Change 2014: Impacts, Adaptation, and*
21 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth*
22 *Assessment Report of the Intergovernmental Panel on Climate Change*, C.B. Field et al., Eds.,
23 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 271–359.
- 24 Sheehy, T., F. Kolahdooz, C. Roache, and S. Sharma, 2015: Traditional food consumption is associated
25 with better diet quality and adequacy among Inuit adults in Nunavut, Canada. *Int. J. Food Sci. Nutr.*,
26 **66**, 445–451, doi:10.3109/09637486.2015.1035232.
27 <http://www.tandfonline.com/doi/full/10.3109/09637486.2015.1035232> (Accessed April 18, 2019).
- 28 Shi, S., W. Zhang, P. Zhang, Y. Yu, and F. Ding, 2013: A synthesis of change in deep soil organic carbon
29 stores with afforestation of agricultural soils. *For. Ecol. Manage.*, **296**, 53–63,
30 doi:10.1016/j.foreco.2013.01.026.
- 31 Shimamoto, C. Y., A. A. Padial, C. M. Da Rosa, and M. C. M. M. Marques, 2018: Restoration of
32 ecosystem services in tropical forests: A global meta-analysis. *PLoS One*, **13**, 1–16,
33 doi:10.1371/journal.pone.0208523.
- 34 Shoyama, K., 2008: Reforestation of abandoned pasture on Hokkaido, northern Japan: Effect of
35 plantations on the recovery of conifer-broadleaved mixed forest. *Landsc. Ecol. Eng.*, **4**, 11–23,
36 doi:10.1007/s11355-008-0034-7.
- 37 Shtienberg, D., 2013: Will Decision-Support Systems Be Widely Used for the Management of Plant
38 Diseases? *Annu. Rev. Phytopathol.*, **51**, 1–16.
- 39 Siebert, S., M. Kummu, M. Porkka, P. Döll, N. Ramankutty, and B. R. Scanlon, 2015: A global data set
40 of the extent of irrigated land from 1900 to 2005. *Hydrol. Earth Syst. Sci.*, doi:10.5194/hess-19-
41 1521-2015.
- 42 Silveira, L., P. Gamazo, J. Alonso, and L. Martínez, 2016: Effects of afforestation on groundwater
43 recharge and water budgets in the western region of Uruguay. *Hydrol. Process.*, **30**, 3596–3608,
44 doi:10.1002/hyp.10952.

- 1 Sivakumar, M. V. K., 2007: Interactions between climate and desertification. *Agric. For. Meteorol.*,
2 doi:10.1016/j.agrformet.2006.03.025.
- 3 Sloan, S., and J. A. Sayer, 2015: Forest Resources Assessment of 2015 shows positive global trends but
4 forest loss and degradation persist in poor tropical countries. *For. Ecol. Manage.*, **352**, 134–145,
5 doi:10.1016/j.foreco.2015.06.013.
- 6 Smith, M. D., M. P. Rabbitt, and A. Coleman-Jensen, 2017: Who are the World's Food Insecure? New
7 Evidence from the Food and Agriculture Organization's Food Insecurity Experience Scale. *World*
8 *Dev.*, **93**, 402–412, doi:10.1016/J.WORLDDEV.2017.01.006.
9 <https://www.sciencedirect.com/science/article/pii/S0305750X17300086> (Accessed May 25, 2018).
- 10 Smith, P., 2016: Soil carbon sequestration and biochar as negative emission technologies. *Glob. Chang.*
11 *Biol.*, **22**, 1315–1324, doi:10.1111/gcb.13178.
- 12 —, and P. J. Gregory, 2013: Climate change and sustainable food production. *Proceedings of the*
13 *Nutrition Society*, Vol. 72 of, 21–28 [https://www.cambridge.org/core/services/aop-cambridge-](https://www.cambridge.org/core/services/aop-cambridge-core/content/view/DE02043AE462DF7F91D88FD4349D38E7/S0029665112002832a.pdf/climate_change_and_sustainable_food_production.pdf)
14 [core/content/view/DE02043AE462DF7F91D88FD4349D38E7/S0029665112002832a.pdf/climate_c](https://www.cambridge.org/core/content/view/DE02043AE462DF7F91D88FD4349D38E7/S0029665112002832a.pdf/climate_change_and_sustainable_food_production.pdf)
15 [hange_and_sustainable_food_production.pdf](https://www.cambridge.org/core/content/view/DE02043AE462DF7F91D88FD4349D38E7/S0029665112002832a.pdf/climate_change_and_sustainable_food_production.pdf) (Accessed January 2, 2018).
- 16 Smith, P., and Coauthors, 2014: Agriculture, Forestry and Other Land Use (AFOLU). *Climate Change*
17 *2014: Mitigation of Climate Change. Contribution of Working Group III to the Fifth Assessment*
18 *Report of the Intergovernmental Panel on Climate Change*, O. Edenhofer et al., Eds., Cambridge
19 University Press, Cambridge, United Kingdom and New York, NY, USA.
- 20 Smith, P., and Coauthors, 2016: Biophysical and economic limits to negative CO₂ emissions. *Nat. Clim.*
21 *Chang.*, **6**, 42–50, doi:DOI: 10.1038/NCLIMATE2870.
- 22 Song, X.-P., 2018: Global Estimates of Ecosystem Service Value and Change: Taking Into Account
23 Uncertainties in Satellite-based Land Cover Data. *Ecol. Econ.*, **143**, 227–235,
24 doi:10.1016/j.ecolecon.2017.07.019.
25 <http://linkinghub.elsevier.com/retrieve/pii/S092180091631309X>.
- 26 —, M. C. Hansen, S. V. Stehman, P. V. Potapov, A. Tyukavina, E. F. Vermote, and J. R. Townshend,
27 2018: Global land change from 1982 to 2016. *Nature*, doi:10.1038/s41586-018-0411-9.
28 <http://www.nature.com/articles/s41586-018-0411-9>.
- 29 Spennemann, P. C., M. Salvia, R. C. Ruscica, A. A. Sörensson, F. Grings, and H. Karszenbaum, 2018:
30 Land-atmosphere interaction patterns in southeastern South America using satellite products and
31 climate models. *Int. J. Appl. Earth Obs. Geoinf.*, **64**, 96–103, doi:10.1016/j.jag.2017.08.016.
- 32 Springmann, M., and Coauthors, 2018: Options for keeping the food system within environmental limits.
33 *Nature*, **562**, 1, doi:10.1038/s41586-018-0594-0. <http://www.nature.com/articles/s41586-018-0594-0>
34 (Accessed October 20, 2018).
- 35 Ssmith, P., and Coauthors, 2013: How much land-based greenhouse gas mitigation can be achieved
36 without compromising food security and environmental goals? *Glob. Chang. Biol.*, **19**, 2285–2302,
37 doi:10.1111/gcb.12160. <http://doi.wiley.com/10.1111/gcb.12160> (Accessed October 9, 2018).
- 38 Stadler, K., and Coauthors, 2018: EXIOBASE 3: Developing a Time Series of Detailed Environmentally
39 Extended Multi-Regional Input-Output Tables. *Journal of Industrial Ecology*.
- 40 Stavi, I., and R. Lal, 2015: Achieving Zero Net Land Degradation: Challenges and opportunities. *J. Arid*
41 *Environ.*, doi:10.1016/j.jaridenv.2014.01.016.
- 42 —, G. Bel, and E. Zaady, 2016: Soil functions and ecosystem services in conventional, conservation,
43 and integrated agricultural systems. A review. *Agron. Sustain. Dev.*, **36**, 32, doi:10.1007/s13593-
44 016-0368-8. <https://doi.org/10.1007/s13593-016-0368-8>.

- 1 Sterling, E., and Coauthors, 2017: Culturally Grounded Indicators of Resilience in Social-Ecological
2 Systems. *Environ. Soc.*, **8**, 63–95.
- 3 Sterner, T., 2003: *Policy instruments for environmental and natural resource management*. Resources for
4 the Future Press, Washington DC, USA, 504 pp.
- 5 Stocker, T. F., and Coauthors, 2013a: Summary for Policymakers. *Climate Change 2013: The Physical
6 Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the
7 Intergovernmental Panel on Climate Change*, 1–29.
- 8 Stocker, T. F., and Coauthors, 2013b: *Technical Summary*. In: *Climate Change 2013: The Physical
9 Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the
10 Intergovernmental Panel on Climate Change*. T.F. Stocker et al., Eds. Cambridge University Press,
11 Cambridge, United Kingdom and New York, NY, USA, 33–115 pp.
- 12 Stockmann, U., and Coauthors, 2013: The knowns, known unknowns and unknowns of sequestration of
13 soil organic carbon. *Agric. Ecosyst. Environ.*, **164**, 80–99, doi:10.1016/J.AGEE.2012.10.001.
14 <https://www.sciencedirect.com/science/article/pii/S0167880912003635> (Accessed May 24, 2018).
- 15 Strack, M., 2008: *Peatland and Climate Change*. M. STRACK, Ed. International Peat Society, Jyväskylä,
16 223 pp.
- 17 Strassburg, B. B. N., and Coauthors, 2017: Moment of truth for the Cerrado hotspot. *Nat. Ecol. Evol.*,
18 doi:10.1038/s41559-017-0099.
- 19 Sunil, N., and S. R. Pandravada, 2015: Alien Crop Resources and Underutilized Species for Food and
20 Nutritional Security of India. *Plant Biology and Biotechnology*, Springer India, New Delhi, 757–775
21 http://link.springer.com/10.1007/978-81-322-2286-6_31 (Accessed January 19, 2018).
- 22 Surminski, S., and D. Oramas-Dorta, 2014: Flood insurance schemes and climate adaptation in
23 developing countries. *Int. J. Disaster Risk Reduct.*, **7**, 154–164,
24 doi:<https://doi.org/10.1016/j.ijdr.2013.10.005>.
25 <http://www.sciencedirect.com/science/article/pii/S2212420913000563>.
- 26 Sutton, P. C., S. J. Anderson, R. Costanza, and I. Kubiszewski, 2016: The ecological economics of land
27 degradation: Impacts on ecosystem service values. *Ecol. Econ.*, **129**, 182–192,
28 doi:10.1016/j.ecolecon.2016.06.016.
- 29 Swain, M., L. Blomqvist, J. McNamara, and W. J. Ripple, 2018: Reducing the environmental impact of
30 global diets. *Sci. Total Environ.*, **610–611**, 1207–1209, doi:10.1016/J.SCITOTENV.2017.08.125.
31 <https://www.sciencedirect.com/science/article/pii/S004896971732123X> (Accessed April 3, 2019).
- 32 Swart, R. O. B., and F. Raes, 2007: Making integration of adaptation and mitigation work: mainstreaming
33 into sustainable development policies? *Clim. Policy*, **7**, 288–303,
34 doi:10.1080/14693062.2007.9685657.
35 <http://www.tandfonline.com/doi/abs/10.1080/14693062.2007.9685657> (Accessed April 3, 2019).
- 36 Tal, A., 2010: Desertification. *The Turning Points of Environmental History*, 146–161.
- 37 Terraube, J., A. Fernandez-Llamazares, and M. Cabeza, 2017: The role of protected areas in supporting
38 human health: a call to broaden the assessment of conservation outcomes. *Curr. Opin. Environ.
39 Sustain.*, **25**, 50–58.
- 40 Theriault, V., M. Smale, and H. Haider, 2017: How Does Gender Affect Sustainable Intensification of
41 Cereal Production in the West African Sahel? Evidence from Burkina Faso. *World Dev.*, **92**, 177–
42 191, doi:10.1016/J.WORLDDEV.2016.12.003.
43 <https://www.sciencedirect.com/science/article/pii/S0305750X16305575> (Accessed April 11, 2019).

- 1 Thompson-Hall, M., E. R. Carr, and U. Pascual, 2016: Enhancing and expanding intersectional research
2 for climate change adaptation in agrarian settings. *Ambio*, **45**, 373–382, doi:10.1007/s13280-016-
3 0827-0. <http://link.springer.com/10.1007/s13280-016-0827-0> (Accessed April 17, 2019).
- 4 Thompson, I. D., K. Okabe, J. A. Parrotta, E. Brockerhoff, H. Jactel, D. I. Forrester, and H. Taki, 2014:
5 Biodiversity and ecosystem services: lessons from nature to improve management of planted forests
6 for REDD-plus. *Biodivers. Conserv.*, **23**, 2613–2635, doi:10.1007/s10531-014-0736-0.
7 <http://link.springer.com/10.1007/s10531-014-0736-0> (Accessed September 20, 2018).
- 8 Thyberg, K. L., and D. J. Tonjes, 2016: Drivers of food waste and their implications for sustainable
9 policy development. *Resour. Conserv. Recycl.*, **106**, 110–123,
10 doi:10.1016/J.RESCONREC.2015.11.016.
11 <https://www.sciencedirect.com/science/article/pii/S0921344915301439> (Accessed May 25, 2018).
- 12 Tian, H., and Coauthors, 2019: Global soil nitrous oxide emissions since the preindustrial era estimated
13 by an ensemble of terrestrial biosphere models: Magnitude, attribution, and uncertainty. *Glob.*
14 *Chang. Biol.*, **25**, 640–659, doi:10.1111/gcb.14514.
15 <https://onlinelibrary.wiley.com/doi/abs/10.1111/gcb.14514>.
- 16 Tigchelaar, M., D. S. Battisti, R. L. Naylor, and D. K. Ray, 2018: Future warming increases probability of
17 globally synchronized maize production shocks. *Proc. Natl. Acad. Sci.*, **115**, 6644–6649,
18 doi:10.1073/pnas.1718031115. <https://www.pnas.org/content/115/26/6644> (Accessed April 3,
19 2019).
- 20 Tilman, D., and M. Clark, 2014: Global diets link environmental sustainability and human health. *Nature*,
21 **515**, 518–522, doi:10.1038/nature13959. <http://www.nature.com/articles/nature13959> (Accessed
22 January 19, 2018).
- 23 Tilman, D., C. Balzer, J. Hill, and B. L. Befort, 2011: Global food demand and the sustainable
24 intensification of agriculture. *Proc. Natl. Acad. Sci.*, **108**, 20260–20264,
25 doi:10.1073/pnas.1116437108.
- 26 Tom Veldkamp, Nico Polman, Stijn Reinhard, M. S., 2011: From scaling to governance of the land
27 system: bridging ecological and economic perspectives. *Ecol. Soc.*, **16**, 1.
28 <http://www.ecologyandsociety.org/vol16/iss1/art1/>.
- 29 Tubiello, F. N., and Coauthors, 2015: The Contribution of Agriculture, Forestry and other Land Use
30 activities to Global Warming, 1990–2012. *Glob. Chang. Biol.*, **21**, 2655–2660,
31 doi:10.1111/gcb.12865. <http://dx.doi.org/10.1111/gcb.12865>.
- 32 Turner, P. A., C. B. Field, D. B. Lobell, D. L. Sanchez, and K. J. Mach, 2018: *Unprecedented rates of*
33 *land-use transformation in modelled climate change mitigation pathways*.
- 34 UN, 2015: *Transforming our world: The 2030 agenda for sustainable development*. New York, USA,
35 USA, 41 pp. [https://sustainabledevelopment.un.org/content/documents/7891Transforming Our
36 World.pdf](https://sustainabledevelopment.un.org/content/documents/7891Transforming%20Our%20World.pdf).
- 37 UNCCD, 2014: *Desertification: The Invisible Frontline*.
38 [http://www.droughtmanagement.info/literature/UNCCD_desertification_the_invisible_frontline_20
39 14.pdf](http://www.droughtmanagement.info/literature/UNCCD_desertification_the_invisible_frontline_2014.pdf).
- 40 UNEP, 2016: “*Global Gender and Environment Outlook*.” Nairobi, Kenya,.
- 41 United Nations, Department of Economic and Social Affairs, P. D., 2017: *World Population Prospects:*
42 *The 2017 Revision, DVD Edition*.

- 1 United Nations, 2018: *2018 Revision of World Urbanization Prospects*.
2 [https://www.un.org/development/desa/publications/2018-revision-of-world-urbanization-](https://www.un.org/development/desa/publications/2018-revision-of-world-urbanization-prospects.html)
3 [prospects.html](https://www.un.org/development/desa/publications/2018-revision-of-world-urbanization-prospects.html).
- 4 Urban, M. C., and Coauthors, 2016: Improving the forecast for biodiversity under climate change.
5 *Science*, **353**, doi:10.1126/science.aad8466.
- 6 USDA, 2007: *Precision Agriculture: NRCS Support for Emerging Technologies*.
7 http://www.nrcs.usda.gov/Internet/FSE_DOCUMENTS/stelprdb1043474.pdf.
- 8 Vadell, E., S. De-Miguel, and J. Pemán, 2016: Large-scale reforestation and afforestation policy in Spain:
9 A historical review of its underlying ecological, socioeconomic and political dynamics. *Land use*
10 *policy*, **55**, 37–48, doi:10.1016/J.LANDUSEPOL.2016.03.017.
11 <https://www.sciencedirect.com/science/article/pii/S0264837716302344> (Accessed September 12,
12 2018).
- 13 Valayamkunnath, P., V. Sridhar, W. Zhao, and R. G. Allen, 2018: Intercomparison of surface energy
14 fluxes, soil moisture, and evapotranspiration from eddy covariance, large-aperture scintillometer,
15 and modeling across three ecosystems in a semiarid climate. *Agric. For. Meteorol.*, **248**, 22–47,
16 doi:10.1016/j.agrformet.2017.08.025. <http://dx.doi.org/10.1016/j.agrformet.2017.08.025>.
- 17 Valentin, C., and Coauthors, 2008: Agriculture , Ecosystems and Environment Runoff and sediment
18 losses from 27 upland catchments in Southeast Asia : Impact of rapid land use changes and
19 conservation practices. *Agric. Ecosyst. Environ.*, **128**, 225–238, doi:10.1016/j.agee.2008.06.004.
- 20 Vaughan, N. E., and C. Gough, 2016: Expert assessment concludes negative emissions scenarios may not
21 deliver. *Environ. Res. Lett.*, **11**, 95003, doi:10.1088/1748-9326/11/9/095003.
22 <http://dx.doi.org/10.1088/1748-9326/11/9/095003>.
- 23 Veldman, J. W., and Coauthors, 2015: Where Tree Planting and Forest Expansion are Bad for
24 Biodiversity and Ecosystem Services. *Bioscience*, **65**, 1011–1018, doi:10.1093/biosci/biv118.
25 [http://academic.oup.com/bioscience/article/65/10/1011/245863/Where-Tree-Planting-and-Forest-](http://academic.oup.com/bioscience/article/65/10/1011/245863/Where-Tree-Planting-and-Forest-Expansion-are-Bad)
26 [Expansion-are-Bad](http://academic.oup.com/bioscience/article/65/10/1011/245863/Where-Tree-Planting-and-Forest-Expansion-are-Bad) (Accessed September 19, 2018).
- 27 Veldman, J. W., F. A. O. Silveira, F. D. Fleischman, N. L. Ascarrunz, and G. Durigan, 2017: Grassy
28 biomes: An inconvenient reality for large-scale forest restoration? A comment on the essay by
29 Chazdon and Laestadius. *Am. J. Bot.*, **104**, 649–651, doi:doi:10.3732/ajb.1600427.
30 <https://onlinelibrary.wiley.com/doi/abs/10.3732/ajb.1600427>.
- 31 Venter, O., and Coauthors, 2016: Sixteen years of change in the global terrestrial human footprint and
32 implications for biodiversity conservation. *Nat. Commun.*, **7**, doi:10.1038/ncomms12558.
- 33 van Vliet, J., D. A. Eitelberg, and P. H. Verburg, 2017: A global analysis of land take in cropland areas
34 and production displacement from urbanization. *Glob. Environ. Chang.*, **43**, 107–115,
35 doi:10.1016/j.gloenvcha.2017.02.001.
- 36 De Vos, J. M., L. N. Joppa, J. L. Gittleman, P. R. Stephens, and S. L. Pimm, 2015: Estimating the normal
37 background rate of species extinction. *Conserv. Biol.*, **29**, 452–462, doi:doi:10.1111/cobi.12380.
38 <https://onlinelibrary.wiley.com/doi/abs/10.1111/cobi.12380>.
- 39 van Vuuren, D. P., and T. R. Carter, 2014: Climate and socio-economic scenarios for climate change
40 research and assessment: reconciling the new with the old. *Clim. Change*, **122**, 415–429,
41 doi:10.1007/s10584-013-0974-2.
- 42 van Vuuren, D. P., and Coauthors, 2017: Energy, land-use and greenhouse gas emissions trajectories
43 under a green growth paradigm. *Glob. Environ. Chang.*, **42**, 237–250,
44 doi:10.1016/J.GLOENVCHA.2016.05.008.
45 <https://www.sciencedirect.com/science/article/pii/S095937801630067X> (Accessed March 22, 2019).

- 1 Vuuren, D. P. Van, and Coauthors, 2018: the need for negative emission technologies. *Nat. Clim. Chang.*,
2 **8**, doi:10.1038/s41558-018-0119-8. <http://dx.doi.org/10.1038/s41558-018-0119-8>.
- 3 Walker, W. E., M. Haasnoot, and J. H. Kwakkel, 2013: Adapt or Perish: A Review of Planning
4 Approaches for Adaptation under Deep Uncertainty. *Sustainability*, **5**, 955–979,
5 doi:10.3390/su5030955. <http://www.mdpi.com/2071-1050/5/3/955>.
- 6 Walters, M., and R. J. Scholes, 2017: *The GEO handbook on biodiversity observation networks*.
7 Springer,.
- 8 Wang, X., A. Biewald, J. P. Dietrich, C. Schmitz, H. Lotze-Campen, F. Humpeöder, B. L. Bodirsky, and
9 A. Popp, 2016: Taking account of governance: Implications for land-use dynamics, food prices, and
10 trade patterns. *Ecol. Econ.*, **122**, 12–24, doi:10.1016/j.ecolecon.2015.11.018.
11 <http://dx.doi.org/10.1016/j.ecolecon.2015.11.018>.
- 12 Wärlind, D., and Coauthors, 2014: Nitrogen feedbacks increase future terrestrial ecosystem carbon uptake
13 in an individual-based dynamic vegetation model. *Biogeosciences*, **11**, 6131–6146, doi:10.5194/bg-
14 11-6131-2014. [https://www.scopus.com/inward/record.uri?eid=2-s2.0-
15 84898401361%7B&%7DpartnerID=40%7B&%7Dmd5=9c68b67b5c217f06f433a5a45e1fec54](https://www.scopus.com/inward/record.uri?eid=2-s2.0-84898401361%7B&%7DpartnerID=40%7B&%7Dmd5=9c68b67b5c217f06f433a5a45e1fec54).
- 16 Warren, D. D. and D. C. and N. R. and J. P. and R., 2014: Global crop yield response to extreme heat
17 stress under multiple climate change futures. *Environ. Res. Lett.*, **9**, 34011.
- 18 Warszawski, L., K. Frieler, V. Huber, F. Piontek, O. Serdeczny, and J. Schewe, 2014: The Inter-Sectoral
19 Impact Model Intercomparison Project (ISI-MIP): Project framework. *Proc. Natl. Acad. Sci.*, **111**,
20 3228–3232, doi:10.1073/pnas.1312330110.
21 <http://www.pnas.org/lookup/doi/10.1073/pnas.1312330110>.
- 22 Watmuff, G., D. J. Reuter, and S. D. Speirs, 2013: Methodologies for assembling and interrogating N, P,
23 K, and S soil test calibrations for Australian cereal, oilseed and pulse crops. *Crop Pasture Sci.*, **64**,
24 424, doi:10.1071/CP12424. <http://www.publish.csiro.au/?paper=CP12424> (Accessed May 21,
25 2018).
- 26 Watson, R. T., I. R. Noble, B. Bolin, N. Ravindranath, D. J. Verardo, and D. J. Dokken, 2000: *Land use,*
27 *land-use change and forestry: a special report of the Intergovernmental Panel on Climate Change*.
28 Cambridge University Press, New York,.
- 29 Weindl, I., and Coauthors, 2017: Livestock and human use of land: Productivity trends and dietary
30 choices as drivers of future land and carbon dynamics. *Glob. Planet. Change*, **159**, 1–10,
31 doi:10.1016/j.gloplacha.2017.10.002.
- 32 Weitzman, M. L., 2014: Can negotiating a uniform carbon price help to internalize the global warming
33 externality? *J. Assoc. Environ. Resour. Econ.*, **1**, 29–49.
- 34 van der Werf, P., and J. A. Gilliland, 2017: A systematic review of food losses and food waste generation
35 in developed countries. *Proc. Inst. Civ. Eng. - Waste Resour. Manag.*, **170**, 66–77,
36 doi:10.1680/jwarm.16.00026. <http://www.icevirtuallibrary.com/doi/10.1680/jwarm.16.00026>
37 (Accessed April 3, 2019).
- 38 West, T. A. P., 2016: Indigenous community benefits from a de-centralized approach to REDD+ in
39 Brazil. *Clim. Policy*, **16**, 924–939, doi:10.1080/14693062.2015.1058238.
40 <https://doi.org/10.1080/14693062.2015.1058238>.
- 41 Wichelns, D., 2017: The water-energy-food nexus: Is the increasing attention warranted, from either a
42 research or policy perspective? *Environ. Sci. Policy*, **69**, 113–123,
43 doi:10.1016/J.ENVSCI.2016.12.018.
44 <https://www.sciencedirect.com/science/article/pii/S1462901116302970> (Accessed April 18, 2019).

- 1 Widener, M. J., L. Minaker, S. Farber, J. Allen, B. Vitali, P. C. Coleman, and B. Cook, 2017: How do
2 changes in the daily food and transportation environments affect grocery store accessibility? *Appl.*
3 *Geogr.*, **83**, 46–62, doi:10.1016/J.APGEOG.2017.03.018.
4 <https://www.sciencedirect.com/science/article/pii/S0143622816303721> (Accessed April 3, 2019).
- 5 Wiedmann, T., and M. Lenzen, 2018: Environmental and social footprints of international trade. *Nat.*
6 *Geosci.*, **11**, 314–321, doi:10.1038/s41561-018-0113-9. [http://www.nature.com/articles/s41561-018-](http://www.nature.com/articles/s41561-018-0113-9)
7 [0113-9](http://www.nature.com/articles/s41561-018-0113-9) (Accessed April 3, 2019).
- 8 Williamson, P., 2016: Scrutinize CO2 removal methods. *Nature*, **530**, 153–155.
- 9 Wilson, S. J., J. Schelhas, R. Grau, A. S. Nanni, and S. Sloan, 2017: Forest ecosystem-service transitions:
10 The ecological dimensions of the forest transition. *Ecol. Soc.*, **22**, doi:10.5751/es-09615-220438.
- 11 Wilting, H. C., A. M. Schipper, M. Bakkenes, J. R. Meijer, and M. A. J. Huijbregts, 2017: Quantifying
12 Biodiversity Losses Due to Human Consumption: A Global-Scale Footprint Analysis. *Environ. Sci.*
13 *Technol.*, **51**, 3298–3306, doi:10.1021/acs.est.6b05296.
- 14 Wise, R. M., I. Fazey, M. S. Smith, S. E. Park, H. C. Eakin, E. R. M. A. Van Garderen, and B. Campbell,
15 2014: Reconceptualising adaptation to climate change as part of pathways of change and response.
16 *Glob. Environ. Chang.*, **28**, 325–336. <https://doi.org/10.1016/j.gloenvcha.2013.12.002>.
- 17 Wisser, D., S. Frolking, E. M. Douglas, B. M. Fekete, C. J. Vörösmarty, and A. H. Schumann, 2008:
18 Global irrigation water demand: Variability and uncertainties arising from agricultural and climate
19 data sets. *Geophys. Res. Lett.*, doi:10.1029/2008GL035296.
- 20 Wolff, S., E. A. Schrammeijer, C. Schulp, and P. H. Verburg, 2018: Meeting global land restoration and
21 protection targets: what would the world look like in 2050? *Glob. Environ. Chang.*, **52**, 259–272,
22 doi:10.1016/j.gloenvcha.2018.08.002.
- 23 Wood, S. A., M. R. Smith, J. Fanzo, R. Remans, and R. S. DeFries, 2018: Trade and the equitability of
24 global food nutrient distribution. *Nat. Sustain.*, **1**, 34–37, doi:10.1038/s41893-017-0008-6.
25 <http://www.nature.com/articles/s41893-017-0008-6> (Accessed April 22, 2018).
- 26 Wu, X., Y. Lu, S. Zhou, L. Chen, and B. Xu, 2016: Impact of climate change on human infectious
27 diseases: Empirical evidence and human adaptation. *Environ. Int.*, **86**, 14–23,
28 doi:10.1016/J.ENVINT.2015.09.007.
29 <https://www.sciencedirect.com/science/article/pii/S0160412015300489> (Accessed May 25, 2018).
- 30 Wunder, S., 2015: Revisiting the concept of payments for environmental services. *Ecol. Econ.*, **117**, 234–
31 243, doi:10.1016/J.ECOLECON.2014.08.016.
32 <https://www.sciencedirect.com/science/article/pii/S0921800914002961> (Accessed November 2,
33 2018).
- 34 Wynes, S., and K. A. Nicholas, 2017: The climate mitigation gap: education and government
35 recommendations miss the most effective individual actions. *Environ. Res. Lett.*, **12**, 74024.
36 <http://stacks.iop.org/1748-9326/12/i=7/a=074024>.
- 37 Xu, Y., 2018: Political economy of land grabbing inside China involving foreign investors. *Third World*
38 *Q.*, **0**, 1–16, doi:10.1080/01436597.2018.1447372. <https://doi.org/10.1080/01436597.2018.1447372>.
- 39 Xue, L., and Coauthors, 2017: Missing Food, Missing Data? A Critical Review of Global Food Losses
40 and Food Waste Data. *Environ. Sci. Technol.*, **51**, 6618–6633, doi:10.1021/acs.est.7b00401.
41 <https://doi.org/10.1021/acs.est.7b00401>.
- 42 Yang, L., L. Chen, W. Wei, Y. Yu, and H. Zhang, 2014: Comparison of deep soil moisture in two re-
43 vegetation watersheds in semi-arid regions. *J. Hydrol.*, **513**, 314–321,
44 doi:10.1016/j.jhydrol.2014.03.049. <http://dx.doi.org/10.1016/j.jhydrol.2014.03.049>.

- 1 Yang, Y., D. Tilman, G. Furey, and C. Lehman, 2019: Soil carbon sequestration accelerated by
2 restoration of grassland biodiversity. *Nat. Commun.*, **10**, 718, doi:10.1038/s41467-019-08636-w.
3 <http://www.nature.com/articles/s41467-019-08636-w> (Accessed March 21, 2019).
- 4 Yirdaw, E., M. Tigabu, and A. Monge, 2017: Rehabilitation of degraded dryland ecosystems – review.
5 *Silva Fenn.*, **51**, doi:10.14214/sf.1673. <http://www.silvafennica.fi/article/1673> (Accessed April 3,
6 2019).
- 7 Yohe, G. W., 2001: Mitigative Capacity – the Mirror Image of Adaptive Capacity on the Emissions Side.
8 *Clim. Change*, **49**, 247–262, doi:10.1023/A:1010677916703.
9 <http://link.springer.com/10.1023/A:1010677916703> (Accessed April 11, 2019).
- 10 Yu, L., and Coauthors, 2014: Meta-discoveries from a synthesis of satellite-based land-cover mapping
11 research. *Int. J. Remote Sens.*, **35**, 4573–4588, doi:10.1080/01431161.2014.930206.
- 12 Yu, Y., K. Feng, and K. Hubacek, 2013: Tele-connecting local consumption to global land use. *Glob.*
13 *Environ. Chang.*, **23**, 1178–1186, doi:10.1016/J.GLOENVCHA.2013.04.006.
14 <https://www.sciencedirect.com/science/article/pii/S0959378013000721> (Accessed April 3, 2019).
- 15 Zaloumis, N. P., and W. J. Bond, 2015: Reforestation of afforestation? the attributes of old growth
16 grasslands in South Africa. *Philos. Trans. R. Soc. B*, **371**, 1–9, doi:10.1098/rstb.2015.0310.
- 17 Zhang, M., and Coauthors, 2014: Response of surface air temperature to small-scale land clearing across
18 latitudes. *Environ. Res. Lett.*, **9**, 34002. <http://stacks.iop.org/1748-9326/9/i=3/a=034002>.
- 19 Zhang, X., E. A. Davidson, D. L. Mauzerall, T. D. Searchinger, P. Dumas, and Y. Shen, 2015: Managing
20 nitrogen for sustainable development. *Nature*, **528**, 51–59, doi:10.1038/nature15743.
- 21 Zhao, L., A. Dai, and B. Dong, 2018: Changes in global vegetation activity and its driving factors during
22 1982–2013. *Agric. For. Meteorol.*, doi:10.1016/j.agrformet.2017.11.013.
- 23 Zheng, H., Y. Wang, Y. Chen, and T. Zhao, 2016: Effects of large-scale afforestation project on the
24 ecosystem water balance in humid areas: An example for southern China. *Ecol. Eng.*, **89**, 103–108,
25 doi:10.1016/j.ecoleng.2016.01.013. <http://dx.doi.org/10.1016/j.ecoleng.2016.01.013>.
- 26 Zhu, Z., and Coauthors, 2016: Greening of the Earth and its drivers. *Nat. Clim. Chang.*, **6**, 791–795,
27 doi:10.1038/nclimate3004. <http://www.nature.com/articles/nclimate3004> (Accessed August 9,
28 2018).
- 29 Ziadat, F., S. Bunning, S. Corsi, and R. Vargas, 2018: Sustainable soil and land management for climate
30 smart agriculture. *Climate Smart Agriculture Sourcebook*, Food and Agriculture Organization of the
31 United Nations [http://www.fao.org/climate-smart-agriculture-sourcebook/production-
32 resources/module-b7-soil/b7-overview/en/?type=111](http://www.fao.org/climate-smart-agriculture-sourcebook/production-resources/module-b7-soil/b7-overview/en/?type=111).
- 33 Ziese, M., and Coauthors, 2014: The GPCC Drought Index - A new, combined and gridded global
34 drought index. *Earth Syst. Sci. Data*, **6**, 285–295, doi:10.5194/essd-6-285-2014.
- 35 Ziska, L. H., and Coauthors, 2016: Rising atmospheric CO₂ is reducing the protein concentration of a
36 floral pollen source essential for North American bees. *Proceedings. Biol. Sci.*, **283**, 20160414,
37 doi:10.1098/rspb.2016.0414. <http://www.ncbi.nlm.nih.gov/pubmed/27075256> (Accessed May 25,
38 2018).
- 39 Zorya, Sergiy, Morgan, Nancy, Diaz Rios, Luz, Hodges, Rick, Bennett, Ben, Stathers, Tanya, Mwebaze,
40 Paul and Lamb, J., 2011: *Missing food: the case of postharvest grain losses in sub-Saharan Africa*.
41 Washington DC, USA,
42 http://siteresources.worldbank.org/INTARD/Resources/MissingFoods10_web_final1.pdf.

43

44

1

2 **Supplementary Material**3 **Table SM.1.1 Observations related to variables indicative of land management, and their**
4 **uncertainties**

| LM-related process | Observations methodology | Scale of observations (space and time) | Uncertainties ² | Pros and cons | Select literature |
|--------------------|---|--|----------------------------|---|--|
| GHG emissions | Micrometeorological fluxes (CO ₂) | 1-10 ha 0.5hr- >10 y | 5-15% | <u>Pros</u> Larger footprints Continuous monitoring Less disturbance on monitored system | (Richardson et al. 2006; Luysaert et al. 2007; Foken and Napo 2008; Mauder et al. 2013; Peltola et al. 2014; Wang et al. 2015; |
| | Micrometeorological fluxes (CH ₄) | | 10-40% | Detailed protocols | Rannik et al. 2015; |
| | Micrometeorological fluxes (N ₂ O) | | 20-50% | <u>Cons</u> Limitations by fetch and turbulence scale Not all trace gases | Campioli et al. 2016; Rannik et al. 2016; Wang et al. 2017a; Brown and Wagner-Riddle 2017; Desjardins et al. 2018) |
| | Soil chambers (CO ₂) | 0.01-1 ha 0.5hr - 1 y | 5%-15% | <u>Pros</u> Relatively inexpensive | (Vargas and Allen 2008; Lavoie et al. 2015; Barton et al. 2015; Dossa et al. 2015; |
| | Soil chambers (CH ₄) | | 5% - 25% | Possibility of manipulation experiments | Ogle et al. 2016; |
| | Soil chambers (N ₂ O) | | 53%- 100% ³ | Large range of trace gases | Pirk et al. 2016; Morin et al. 2017; |
| | | | | <u>Cons</u> Smaller footprint Complicate upscaling Static pressure interference | Lammirato et al. 2018) |
| | Atmospheric inversions (CO ₂) | Regional 1->10 y | 50% | <u>Pros</u> Integration on large scale | (Wang et al. 2017b) |
| | Atmospheric inversions | | 3-8% | Attribution detection (with | (Pison et al. 2018) |

² FOOTNOTE: Uncertainty here is defined as the coefficient of variation CV. In the case of micrometeorological fluxes they refer to random errors and CV of daily average

³ FOOTNOTE: > 100 for fluxes less than 5g N₂O-N ha⁻¹ d⁻¹

| (CH ₄) | | | | 14C) | |
|--------------------|---|----------------------------------|----------|--|---|
| Carbon balance | Soil carbon point measurements | 0.01ha-1ha >5 y | 5-20% | <p>Rigorously derived uncertainty</p> <p><u>Cons</u></p> <p>Not suited at farm scale</p> <p>Large high precision observation network required</p> <p><u>Pros</u></p> <p>Easy protocol</p> <p>Well established analytics</p> <p><u>Cons</u></p> <p>Need high number of samples for upscaling</p> <p>Detection limit is high</p> | (Chiti et al. 2018; Castaldi et al. 2018; Chen et al. 2018; Deng et al. 2018) |
| | Biomass measurements | 0.01ha – 1ha 1-5 y | 2-8% | <p><u>Pros</u></p> <p>Well established allometric equations</p> <p>High accuracy at plot level</p> <p><u>Cons</u></p> <p>Difficult to scale up</p> <p>Labour intensive</p> | (Pelletier et al. 2012; Henry et al. 2015; Vanguelova et al. 2016; Djomo et al. 2016; Forrester et al. 2017; Xu et al. 2017; Marziliano et al. 2017; Clark et al. 2017; Disney et al. 2018; Urbazaev et al. 2018; Paul et al. 2018) |
| Water balance | Soil moisture (IoT sensors, Cosmic rays, Thermo-optical sensing etc.) | 0.01ha – regional 0.5hr- <1y | 3-5% vol | <p><u>Pros</u></p> <p>New technology</p> <p>Big data analytics</p> <p>Relatively inexpensive</p> <p><u>Cons</u></p> <p>Scaling problems</p> | (Yu et al. 2013; Zhang and Zhou 2016; Iwata et al. 2017; McJannet et al. 2017; Karthikeyan et al. 2017; Iwata et al. 2017; Cao et al. 2018; Amaral et al. 2018; Moradizadeh and Saradjian 2018; Strati et al. 2018) |
| | Evapotranspiration | 0.01ha – Regional 0.5hr- >10y | 10-20% | <p><u>Pros</u></p> <p>Well established methods</p> <p>Easy integration in models and DSS</p> <p><u>Cons</u></p> <p>Partition of fluxes need additional measurements</p> | (Zhang et al. 2017; Papadimitriou et al. 2017; Kaushal et al. 2017; Valayamkunnath et al. 2018; Valayamkunnath et al. 2018; Tie et al. 2018; Wang et al. 2018) |
| Soil Erosion | Sediment transport | 1 ha – Regional 1d - >10y | -21-34% | <p><u>Pros</u></p> <p>Long history of</p> | (Efthimiou 2018; García-Barrón et al. |

| | | | | | |
|------------|-----------|-----------------------------------|-----------|----------------------|------------------------|
| | | | | methods | 2018; Fiener et al. |
| | | | | Integrative tools | 2018) |
| | | | | <u>Cons</u> | |
| | | | | Validation is | |
| | | | | lacking | |
| | | | | Labour intensive | |
| Land cover | Satellite | 0.01ha – Regional 1d - >10y | 16 - 100% | <u>Pros</u> | (Olofsson et al. 2014; |
| | | | | Increasing platforms | Liu et al. 2018; Yang |
| | | | | available | et al. 2018) |
| | | | | Consolidated | |
| | | | | algorithms | |
| | | | | <u>Cons</u> | |
| | | | | Need validation | |
| | | | | Lack of common | |
| | | | | Land Use | |
| | | | | definitions | |

1

2

3

1 **Table SM. 1.2 Possible uncertainties decision making faces** (following (Hansson and Hadorn 2016))

2

| Type | Knowledge gaps | Understanding the uncertainties |
|--|--|---|
| Uncertainty of consequences | Do the model(s) adequately represent the target system? What are the numerical values of input parameters, boundary conditions, or initial conditions? What are all potential events that we would take into account if we were aware of them? Will future events relevant for our decisions, including expected impacts from these decisions, in fact take place? | Ensemble approaches; downscaling Benchmarking, sensitivity analyses Scenario approaches |
| Moral uncertainty | How to (ethically) evaluate the decisions? What values to base the decision on (→ often unreliable ranking of values not doing justice to the range of values at stake, cp. Sen 1992), including choice of discount rate, risk attitude (risk aversion, risk neutral, ...) Which ethical principles? (i.e. utilitarian, deontic, virtue, or other?) | Possibly scenario analysis Identification of lock-in effects and path-dependency (e.g. Kinsley et al 2016) |
| Uncertainty of demarcation | What are the options that we can actually choose between? (not fully known because “decision costs” may be high, or certain options are not “seen” as they are outside current ideologies). How can the mass of decisions divided into individual decisions? e.g. how this influences international negotiations and the question who does what and when (cp. Hammond et al. 1999). | Possibly scenario analysis |
| Uncertainty of consequences & uncertainty of demarcation | What effects does a decision have when combined with the decision of others? (e.g. other countries may follow the inspiring example in climate reduction of country X, or they use it solely in their own economic interest) | Games |
| Uncertainty of demarcation & moral uncertainty | How would we decide in the future? (Spohn 1977; Rabinowicz 2002) | |

3

4 **References SM1**

- 5 Amaral, A. M., F. R. Cabral Filho, L. M. Vellame, M. B. Teixeira, F. A. L. Soares, and L. N. S. do. dos
6 Santos, 2018: Uncertainty of weight measuring systems applied to weighing lysimeters. *Comput.*
7 *Electron. Agric.*, **145**, 208–216, doi:10.1016/j.compag.2017.12.033.
8 <http://linkinghub.elsevier.com/retrieve/pii/S0168169917301047> (Accessed May 20, 2018).
- 9 Barton, L., B. Wolf, D. Rowlings, C. Scheer, R. Kiese, P. Grace, K. Stefanova, and K. Butterbach-Bahl,
10 2015: Sampling frequency affects estimates of annual nitrous oxide fluxes. *Sci. Rep.*, **5**, 1–9,
11 doi:10.1038/srep15912. <http://dx.doi.org/10.1038/srep15912>.
- 12 Brown, S. E., and C. Wagner-Riddle, 2017: Assessment of random errors in multi-plot nitrous oxide flux
13 gradient measurements. *Agric. For. Meteorol.*, **242**, 10–20, doi:10.1016/j.agrformet.2017.04.005.
14 <http://dx.doi.org/10.1016/j.agrformet.2017.04.005>.
- 15 Campioli, M., and Coauthors, 2016: ARTICLE Evaluating the convergence between eddy-covariance and
16 biometric methods for assessing carbon budgets of forests. *Nat. Commun.*, **7**,

- 1 doi:10.1038/ncomms13717. <https://www.nature.com/articles/ncomms13717.pdf> (Accessed May 19,
2 2018).
- 3 Cao, D.-F., B. Shi, G.-Q. Wei, S.-E. Chen, and H.-H. Zhu, 2018: An improved distributed sensing method
4 for monitoring soil moisture profile using heated carbon fibers. *Meas. J. Int. Meas. Confed.*, **123**,
5 doi:10.1016/j.measurement.2018.03.052.
- 6 Castaldi, F., S. Chabrilat, C. Chartin, V. Genot, A. R. Jones, and B. van Wesemael, 2018: Estimation of
7 soil organic carbon in arable soil in Belgium and Luxembourg with the LUCAS topsoil database.
8 *Eur. J. Soil Sci.*, doi:10.1111/ejss.12553. <http://doi.wiley.com/10.1111/ejss.12553> (Accessed May
9 20, 2018).
- 10 Chen, S., M. P. Martin, N. P. A. Saby, C. Walter, D. A. Angers, and D. Arrouays, 2018: Fine resolution
11 map of top- and subsoil carbon sequestration potential in France. *Sci. Total Environ.*, **630**, 389–400,
12 doi:10.1016/J.SCITOTENV.2018.02.209.
13 <https://www.sciencedirect.com/science/article/pii/S0048969718306089?via%3Dihub> (Accessed
14 May 20, 2018).
- 15 Chiti, T., E. Blasi, G. Pellis, L. Perugini, M. V. Chiriaco, and R. Valentini, 2018: Soil organic carbon
16 pool's contribution to climate change mitigation on marginal land of a Mediterranean montane area
17 in Italy. *J. Environ. Manage.*, **218**, 593–601, doi:10.1016/j.jenvman.2018.04.093.
18 <https://www.sciencedirect.com/science/article/pii/S030147971830481X?via%3Dihub> (Accessed
19 May 20, 2018).
- 20 Clark, D. A., S. Asao, R. Fisher, S. Reed, P. B. Reich, M. G. Ryan, T. E. Wood, and X. Yang, 2017:
21 Reviews and syntheses: Field data to benchmark the carbon cycle models for tropical forests.
22 *Biogeosciences*, **14**, 4663–4690, doi:10.5194/bg-14-4663-2017.
- 23 Deng, X., and Coauthors, 2018: Baseline map of organic carbon stock in farmland topsoil in East China.
24 *Agric. Ecosyst. Environ.*, **254**, 213–223, doi:10.1016/J.AGEE.2017.11.022.
25 <https://www.sciencedirect.com/science/article/pii/S0167880917305194?via%3Dihub> (Accessed
26 May 20, 2018).
- 27 Desjardins, R. L., and Coauthors, 2018: The challenge of reconciling bottom-up agricultural methane
28 emissions inventories with top-down measurements. *Agric. For. Meteorol.*, **248**, 48–59,
29 doi:10.1016/j.agrformet.2017.09.003. <http://dx.doi.org/10.1016/j.agrformet.2017.09.003>.
- 30 Disney, M. I., M. Boni Vicari, A. Burt, K. Calders, S. L. Lewis, P. Raunonen, and P. Wilkes, 2018:
31 Weighing trees with lasers: advances, challenges and opportunities. *Interface Focus*, **8**, 20170048,
32 doi:10.1098/rsfs.2017.0048.
33 <http://rsfs.royalsocietypublishing.org/lookup/doi/10.1098/rsfs.2017.0048> (Accessed May 20, 2018).
- 34 Djomo, A. N., and Coauthors, 2016: Tree allometry for estimation of carbon stocks in African tropical
35 forests. *Forestry*, **89**, 446–455, doi:10.1093/forestry/cpw025.
- 36 Dossa, G. G. O., E. Paudel, H. Wang, K. Cao, D. Schaefer, and R. D. Harrison, 2015: Correct calculation
37 of CO₂ efflux using a closed-chamber linked to a non-dispersive infrared gas analyzer. *Methods
38 Ecol. Evol.*, **6**, 1435–1442, doi:10.1111/2041-210X.12451.
- 39 Efthimiou, N., 2018: The importance of soil data availability on erosion modeling. *CATENA*, **165**, 551–
40 566, doi:10.1016/J.CATENA.2018.03.002.
41 <https://www.sciencedirect.com/science/article/pii/S034181621830078X?via%3Dihub> (Accessed
42 May 20, 2018).
- 43 Fiener, P., and Coauthors, 2018: Uncertainties in assessing tillage erosion – How appropriate are our
44 measuring techniques? *Geomorphology*, **304**, 214–225, doi:10.1016/J.GEOMORPH.2017.12.031.

- 1 <https://www.sciencedirect.com/science/article/pii/S0169555X17305391?via%3Dihub> (Accessed
2 May 20, 2018).
- 3 Foken, T., and C. J. Napo, 2008: *Micrometeorology*. Springer,.
- 4 Forrester, D. I., and Coauthors, 2017: Generalized biomass and leaf area allometric equations for
5 European tree species incorporating stand structure, tree age and climate. *For. Ecol. Manage.*, **396**,
6 160–175, doi:10.1016/j.foreco.2017.04.011. <http://dx.doi.org/10.1016/j.foreco.2017.04.011>.
- 7 García-Barrón, L., J. Morales, and A. Sousa, 2018: A new methodology for estimating rainfall
8 aggressiveness risk based on daily rainfall records for multi-decennial periods. *Sci. Total Environ.*,
9 **615**, 564–571, doi:10.1016/j.scitotenv.2017.09.305. <https://doi.org/10.1016/j.scitotenv.2017.09.305>.
- 10 Hammond, J. S., R. L. Keeney, and H. R., 1999: Smart choices: a practical guide to making better life
11 decisions. *Broadway Books*, New York, USA.
- 12 Hansson, S. O., and G. H. Hadorn, 2016: Introducing the Argumentative Turn in Policy Analysis. *The*
13 *Argumentative Turn in Policy Analysis, Logic, Argumentation & Reasoning*, Springer, Cham, 11–35
14 https://link.springer.com/chapter/10.1007/978-3-319-30549-3_2.
- 15 Henry, M., and Coauthors, 2015: Recommendations for the use of tree models to estimate national forest
16 biomass and assess their uncertainty. *Ann. For. Sci.*, **72**, 769–777, doi:10.1007/s13595-015-0465-x.
- 17 Iwata, Y., T. Miyamoto, K. Kameyama, and M. Nishiya, 2017: Effect of sensor installation on the
18 accurate measurement of soil water content. *Eur. J. Soil Sci.*, **68**, 817–828, doi:10.1111/ejss.12493.
- 19 Karthikeyan, L., M. Pan, N. Wanders, D. N. Kumar, and E. F. Wood, 2017: Four decades of microwave
20 satellite soil moisture observations: Part 1. A review of retrieval algorithms. *Adv. Water Resour.*,
21 **109**, 106–120, doi:10.1016/j.advwatres.2017.09.006.
22 <http://dx.doi.org/10.1016/j.advwatres.2017.09.006>.
- 23 Kaushal, S. S., A. J. Gold, and P. M. Mayer, 2017: Land use, climate, and water resources-global stages
24 of interaction. *Water (Switzerland)*, **9**, 815, doi:10.3390/w9100815.
- 25 Lammirato, C., U. Lebender, J. Tierling, and J. Lammel, 2018: Analysis of uncertainty for N₂O fluxes
26 measured with the closed-chamber method under field conditions: Calculation method, detection
27 limit, and spatial variability. *J. Plant Nutr. Soil Sci.*, **181**, 78–89, doi:10.1002/jpln.201600499.
28 <http://doi.wiley.com/10.1002/jpln.201600499> (Accessed May 20, 2018).
- 29 Lavoie, M., C. L. Phillips, and D. Risk, 2015: A practical approach for uncertainty quantification of high-
30 frequency soil respiration using Forced Diffusion chambers. *J. Geophys. Res. Biogeosciences*, **120**,
31 128–146, doi:10.1002/2014JG002773. <http://doi.wiley.com/10.1002/2014JG002773> (Accessed May
32 20, 2018).
- 33 Liu, X., and Coauthors, 2018: Comparison of country-level cropland areas between ESA-CCI land cover
34 maps and FAOSTAT data. *Int. J. Remote Sens.*, doi:10.1080/01431161.2018.1465613.
- 35 Luyssaert, S., and Coauthors, 2007: CO₂ balance of boreal, temperate, and tropical forests derived from a
36 global database. *Glob. Chang. Biol.*, **13**, 2509–2537, doi:10.1111/j.1365-2486.2007.01439.x.
37 <http://doi.wiley.com/10.1111/j.1365-2486.2007.01439.x> (Accessed May 20, 2018).
- 38 Marziliano, P., G. Menguzzato, and V. Coletta, 2017: Evaluating Carbon Stock Changes in Forest and
39 Related Uncertainty. *Sustainability*, **9**, 1702, doi:10.3390/su9101702. <http://www.mdpi.com/2071-1050/9/10/1702>.
- 40
- 41 Mauder, M., M. Cuntz, C. Drüe, A. Graf, C. Rebmann, H. P. Schmid, M. Schmidt, and R. Steinbrecher,
42 2013: A strategy for quality and uncertainty assessment of long-term eddy-covariance
43 measurements. *Agric. For. Meteorol.*, **169**, 122–135, doi:10.1016/j.agrformet.2012.09.006.
44 <http://linkinghub.elsevier.com/retrieve/pii/S0168192312002808> (Accessed October 11, 2018).

- 1 McJannet, D., A. Hawdon, B. Baker, L. Renzullo, and R. Searle, 2017: Multiscale soil moisture estimates
2 using static and roving cosmic-ray soil moisture sensors. *Hydrol. Earth Syst. Sci. Discuss.*, 1–28,
3 doi:10.5194/hess-2017-358. <https://www.hydrol-earth-syst-sci-discuss.net/hess-2017-358/>.
- 4 Moradizadeh, M., and M. R. Saradjian, 2018: Estimation of improved resolution soil moisture in
5 vegetated areas using passive AMSR-E data. *J. Earth Syst. Sci.*, **127**, 24, doi:10.1007/s12040-018-
6 0925-4. <http://link.springer.com/10.1007/s12040-018-0925-4> (Accessed May 20, 2018).
- 7 Morin, T. H., G. Bohrer, K. C. Stefanik, A. C. Rey-Sanchez, A. M. Matheny, and W. J. Mitsch, 2017:
8 Combining eddy-covariance and chamber measurements to determine the methane budget from a
9 small, heterogeneous urban floodplain wetland park. *Agric. For. Meteorol.*, **237238**, 160–170,
10 doi:10.1016/j.agrformet.2017.01.022. [https://ac.els-cdn.com/S0168192317300321/1-s2.0-
11 S0168192317300321-main.pdf?_tid=c292fa86-eb27-11e7-b381-
12 00000aab0f02&acdnt=1514394346_3c3a4254238776e6cfea1daaf2fd520c](https://ac.els-cdn.com/S0168192317300321/1-s2.0-S0168192317300321-main.pdf?_tid=c292fa86-eb27-11e7-b381-00000aab0f02&acdnt=1514394346_3c3a4254238776e6cfea1daaf2fd520c) (Accessed December
13 27, 2017).
- 14 Ogle, K., E. Ryan, F. A. Dijkstra, and E. Pendall, 2016: Quantifying and reducing uncertainties in
15 estimated soil CO₂ fluxes with hierarchical data-model integration. *J. Geophys. Res.*
16 *Biogeosciences*, **121**, 2935–2948, doi:10.1002/2016JG003385.
17 <http://doi.wiley.com/10.1002/2016JG003385> (Accessed May 20, 2018).
- 18 Olofsson, P., G. M. Foody, M. Herold, S. V. Stehman, C. E. Woodcock, and M. A. Wulder, 2014: Good
19 practices for estimating area and assessing accuracy of land change. *Remote Sens. Environ.*, **148**,
20 42–57, doi:10.1016/j.rse.2014.02.015. [https://ac.els-cdn.com/S0034425714000704/1-s2.0-
21 S0034425714000704-main.pdf?_tid=d8ceec27-1479-4fc8-b527-
22 aca3eaf8d643&acdnt=1526841950_ded0bd23b704a4458fad2c248c6d08d0](https://ac.els-cdn.com/S0034425714000704/1-s2.0-S0034425714000704-main.pdf?_tid=d8ceec27-1479-4fc8-b527-aca3eaf8d643&acdnt=1526841950_ded0bd23b704a4458fad2c248c6d08d0) (Accessed May 20,
23 2018).
- 24 Papadimitriou, L. V., A. G. Koutroulis, M. G. Grillakis, and I. K. Tsanis, 2017: The effect of GCM biases
25 on global runoff simulations of a land surface model. *Hydrol. Earth Syst. Sci.*, **21**, 4379–4401,
26 doi:10.5194/hess-21-4379-2017.
- 27 Paul, K. I., and Coauthors, 2018: Using measured stocks of biomass and litter carbon to constrain
28 modelled estimates of sequestration of soil organic carbon under contrasting mixed-species
29 environmental plantings. *Sci. Total Environ.*, **615**, 348–359, doi:10.1016/j.scitotenv.2017.09.263.
30 <https://doi.org/10.1016/j.scitotenv.2017.09.263>.
- 31 Pelletier, J., K. R. Kirby, and C. Potvin, 2012: Significance of carbon stock uncertainties on emission
32 reductions from deforestation and forest degradation in developing countries. *For. Policy Econ.*, **24**,
33 3–11, doi:10.1016/j.forpol.2010.05.005. <http://dx.doi.org/10.1016/j.forpol.2010.05.005>.
- 34 Peltola, O., and Coauthors, 2014: Evaluating the performance of commonly used gas analysers for
35 methane eddy covariance flux measurements: the InGOS inter-comparison field experiment.
36 *Biogeosciences*, **11**, 3163–3186, doi:10.5194/bg-11-3163-2014.
37 www.biogeosciences.net/11/3163/2014/ (Accessed May 20, 2018).
- 38 Pirk, N., M. Mastepanov, F.-J. W. Parmentier, M. Lund, P. Crill, and T. R. Christensen, 2016:
39 Calculations of automatic chamber flux measurements of methane and carbon dioxide using short
40 time series of concentrations. *Biogeosciences*, **13**, 903–912, doi:10.5194/bg-13-903-2016.
41 www.biogeosciences.net/13/903/2016/ (Accessed May 20, 2018).
- 42 Pison, I., and Coauthors, 2018: How a European network may help with estimating methane emissions on
43 the French national scale. *Atmos. Chem. Phys.*, **185194**, 3779–3798, doi:10.5194/acp-18-3779-2018.
44 <https://www.atmos-chem-phys.net/18/3779/2018/acp-18-3779-2018.pdf> (Accessed May 20, 2018).
- 45 Rabinowicz, W., 2002: Does Practical Deliberation Crowd Out Self-Prediction? *Erkenntnis*, **57**, 91–122,
46 doi:10.1023/A:1020106622032. <https://link.springer.com/article/10.1023/A:1020106622032>.

- 1 Rannik, Ü., O. Peltola, and I. Mammarella, 2016: Random uncertainties of flux measurements by the
2 eddy covariance technique. *Atmos. Meas. Tech.*, **9**, 5163–5181, doi:10.5194/amt-9-5163-2016.
3 www.atmos-meas-tech.net/9/5163/2016/ (Accessed December 27, 2017).
- 4 Spohn, W., 1977: “Where Luce and Krantz do really generalize Savage’s decision model.” *Erkenntnis*,
5 **11**, 113–134. <https://link.springer.com/content/pdf/10.1007/BF00169847.pdf>.
- 6 Strati, V., and Coauthors, 2018: Modelling Soil Water Content in a Tomato Field: Proximal Gamma Ray
7 Spectroscopy and Soil–Crop System Models. *Agriculture*, **8**, 60, doi:10.3390/agriculture8040060.
8 <http://www.mdpi.com/2077-0472/8/4/60> (Accessed May 20, 2018).
- 9 Tie, Q., H. Hu, F. Tian, and N. M. Holbrook, 2018: Comparing different methods for determining forest
10 evapotranspiration and its components at multiple temporal scales. *Sci. Total Environ.*, **633**, 12–29,
11 doi:10.1016/j.scitotenv.2018.03.082.
- 12 Urbazaev, M., C. Thiel, F. Cremer, R. Dubayah, M. Migliavacca, M. Reichstein, and C. Schmullius,
13 2018: Estimation of forest aboveground biomass and uncertainties by integration of field
14 measurements, airborne LiDAR, and SAR and optical satellite data in Mexico. *Carbon Balance*
15 *Manag.*, **13**, doi:10.1186/s13021-018-0093-5. <https://doi.org/10.1186/s13021-018-0093-5> (Accessed
16 May 20, 2018).
- 17 Valayamkunnath, P., V. Sridhar, W. Zhao, and R. G. Allen, 2018: Intercomparison of surface energy
18 fluxes, soil moisture, and evapotranspiration from eddy covariance, large-aperture scintillometer,
19 and modeling across three ecosystems in a semiarid climate. *Agric. For. Meteorol.*, **248**, 22–47,
20 doi:10.1016/j.agrformet.2017.08.025. <http://dx.doi.org/10.1016/j.agrformet.2017.08.025>.
- 21 Vanguelova, E. I., and Coauthors, 2016: Sources of errors and uncertainties in the assessment of forest
22 soil carbon stocks at different scales—review and recommendations. *Environ. Monit. Assess.*, **188**,
23 doi:10.1007/s10661-016-5608-5.
- 24 Vargas, R., and M. F. Allen, 2008: Environmental controls and the influence of vegetation type, fine roots
25 and rhizomorphs on diel and seasonal variation in soil respiration. *New Phytol.*, **179**, 460–471,
26 doi:10.1111/j.1469-8137.2008.02481.x. <http://doi.wiley.com/10.1111/j.1469-8137.2008.02481.x>
27 (Accessed May 20, 2018).
- 28 Wang, E., and Coauthors, 2018: Making sense of cosmic-ray soil moisture measurements and eddy
29 covariance data with regard to crop water use and field water balance. *Agric. Water Manag.*, **204**,
30 271–280, doi:10.1016/J.AGWAT.2018.04.017.
31 <https://www.sciencedirect.com/science/article/pii/S0378377418303871?via%3Dihub> (Accessed
32 May 20, 2018).
- 33 Wang, X., C. Wang, and B. Bond-Lamberty, 2017a: Quantifying and reducing the differences in forest
34 CO₂ fluxes estimated by eddy covariance, biometric and chamber methods: A global synthesis.
35 doi:10.1016/j.agrformet.2017.07.023. www.elsevier.com/locate/agrformet (Accessed December 27,
36 2017).
- 37 Wang, Y., and Coauthors, 2017b: Estimation of observation errors for large-scale atmospheric inversion
38 of CO₂ emissions from fossil fuel combustion. *Tellus B Chem. Phys. Meteorol.*, **69**, 1325723,
39 doi:10.1080/16000889.2017.1325723.
40 <http://www.tandfonline.com/action/journalInformation?journalCode=zclb20> (Accessed May 20,
41 2018).
- 42 Xu, L., and Coauthors, 2017: Spatial Distribution of Carbon Stored in Forests of the Democratic Republic
43 of Congo. *Sci. Rep.*, **7**, 1–12, doi:10.1038/s41598-017-15050-z.
- 44 Yang, L., K. Jia, S. Liang, M. Liu, X. Wei, Y. Yao, X. Zhang, and D. Liu, 2018: Spatio-Temporal
45 Analysis and Uncertainty of Fractional Vegetation Cover Change over Northern China during 2001–

- 1 2012 Based on Multiple Vegetation Data Sets. *Remote Sens.*, **10**, 549, doi:10.3390/rs10040549.
2 <http://www.mdpi.com/2072-4292/10/4/549> (Accessed May 20, 2018).
- 3 Zhang, D., and G. Zhou, 2016: Estimation of Soil Moisture from Optical and Thermal Remote Sensing: A
4 Review. *Sensors*, **16**, 1308, doi:10.3390/s16081308. <http://www.mdpi.com/1424-8220/16/8/1308>.
- 5 Zhang, Y., F. H. S. Chiew, J. Peña-Arancibia, F. Sun, H. Li, and R. Leuning, 2017: Global variation of
6 transpiration and soil evaporation and the role of their major climate drivers. *J. Geophys. Res.*, **122**,
7 6868–6881, doi:10.1002/2017JD027025.
- 8

Chapter 2: Land-Climate Interactions

Coordinating Lead Authors: Gensuo Jia (China), Elena Shevliakova (The United States of America)

Lead Authors: Paulo Artaxo (Brazil), Nathalie De Noblet-Ducoudré (France), Richard Houghton (The United States of America), Joanna House (United Kingdom), Kaoru Kitajima (Japan), Christopher Lennard (South Africa), Alexander Popp (Germany), Andrey Sirin (The Russian Federation), Raman Sukumar (India), Louis Verchot (Colombia/The United States of America)

Contributing Authors: William Anderegg (The United States of America), Edward Armstrong (United Kingdom), Ana Bastos (Portugal/Germany), Terje Koren Bernsten (Norway), Peng Cai (China), Katherine Calvin (The United States of America), Francesco Cherubini (Italy), Sarah Connors (France/United Kingdom), Annette Cowie (Australia), Edouard Davin (Switzerland/ France), Cecile De Klein (New Zealand), Giacomo Grassi (Italy/EU), Rafiq Hamdi (Belgium), Florian Humpenöder (Germany), David Kanter (The United States of America), Gerhard Krinner (France), Sonali McDermid (India/The United States of America), Devaraju Narayanappa (India/France), Josep Peñuelas (Spain), Prajal Pradhan (Nepal), Benjamin Quesada (Colombia), Stephanie Roe (The Philippines/The United States of America), Robert A. Rohde (The United States of America), Martijn Slot (Panama), Rolf Sommer (Germany), Moa Sporre (Norway), Benjamin Sulman (The United States of America), Alasdair Sykes (United Kingdom), Phil Williamson (United Kingdom), Yuyu Zhou (China/The United States of America)

Review Editors: Pierre Bernier (Canada), Jhan Carlo Espinoza (Peru), Sergey Semenov (The Russian Federation)

Chapter Scientist: Xiyang Xu (China)

Date of Draft: 27/4/2019

| | | |
|----|--|------------|
| 1 | Table of Contents | |
| 2 | Chapter 2: Land-Climate Interactions | 1 |
| 3 | Executive Summary..... | 3 |
| 4 | 2.1 Introduction: Land – climate interactions..... | 9 |
| 5 | 2.1.1 Recap of previous IPCC and other relevant reports as baselines..... | 9 |
| 6 | 2.1.2 Introduction to the chapter structure | 11 |
| 7 | Box 2.1: Processes underlying land-climate interactions | 12 |
| 8 | 2.2 The effect of climate variability and change on land | 14 |
| 9 | 2.2.1 Overview of climate impacts on land | 14 |
| 10 | 2.2.2 Climate driven changes in aridity | 16 |
| 11 | 2.2.3 The influence of climate change on food security | 17 |
| 12 | 2.2.4 Climate-driven changes in terrestrial ecosystems | 17 |
| 13 | 2.2.5 Climate extremes and their impact on land functioning | 19 |
| 14 | Cross-Chapter Box 3: Fire and Climate Change | 25 |
| 15 | 2.3 Greenhouse gas fluxes between land and atmosphere..... | 28 |
| 16 | 2.3.1 Carbon Dioxide..... | 29 |
| 17 | 2.3.2 Methane | 36 |
| 18 | 2.3.3 Nitrous Oxide..... | 40 |
| 19 | Box 2.2: Methodologies for estimating national to global scale anthropogenic land carbon fluxes.. | 43 |
| 20 | 2.4 Emissions and impacts of short-lived climate forcers (SLCF) from land | 46 |
| 21 | 2.4.1 Mineral dust | 47 |
| 22 | 2.4.2 Carbonaceous Aerosols..... | 49 |
| 23 | 2.4.3 Biogenic Volatile Organic Compounds (BVOCs)..... | 51 |
| 24 | 2.5 Land impacts on climate and weather through biophysical and GHGs effects | 53 |
| 25 | 2.5.1 Impacts of historical and future anthropogenic land cover changes | 54 |
| 26 | 2.5.2 Impacts of specific land use changes | 60 |
| 27 | 2.5.3 Amplifying / dampening climate changes via land responses | 68 |
| 28 | 2.5.4 Non-local and downwind effects resulting from changes in land cover | 72 |
| 29 | Cross-Chapte Box 4: Climate Change and Urbanisation | 73 |
| 30 | 2.6 Climate consequences of response options | 77 |
| 31 | 2.6.1 Climate impacts of individual response options | 77 |
| 32 | 2.6.2 Integrated pathways for climate change mitigation | 85 |
| 33 | 2.6.3 The contribution of response options to the Paris Agreement | 91 |
| 34 | 2.7 Plant and soil processes underlying land-climate interactions | 94 |
| 35 | 2.7.1 Temperature responses of plant and ecosystem production..... | 94 |
| 36 | 2.7.2 Water transport through soil-plant-atmosphere continuum and drought mortality..... | 95 |
| 37 | 2.7.3 Soil microbial effects on soil nutrient dynamics and plant responses to elevated CO ₂ | 96 |
| 38 | 2.7.4 Vertical distribution of soil organic carbon | 97 |
| 39 | 2.7.5 Soil carbon responses to warming and changes in soil moisture | 97 |
| 40 | 2.7.6 Soil carbon responses to changes in organic-matter inputs by plants | 98 |
| 41 | References | 101 |
| 42 | Appendix | 178 |
| 43 | | |
| 44 | | |
| 45 | | |

1 **Executive Summary**

2
3 Land and climate interact in complex ways through changes in forcing and multiple biophysical and
4 biogeochemical feedbacks across different spatial and temporal scales. This chapter assesses climate impacts
5 on land and land impacts on climate, the human contributions to these changes, as well as land-based
6 adaptation and mitigation response options to combat projected climate changes.
7

8 ***Implications of climate change, variability, and extremes for land systems***

9
10 **It is certain that globally averaged land surface air temperature (LSAT) has risen faster than the**
11 **global mean surface temperature (i.e., combined LSAT and sea surface temperature) from**
12 **preindustrial (1850–1900) to present day (1999–2018). According to the single longest and most**
13 **extensive dataset, the LSAT increase between the preindustrial period and present day was 1.52°C**
14 **(the *very likely* range of 1.39°C to 1.66°C). For the 1880–2018 period, when four independently**
15 **produced datasets exist, the LSAT increase was 1.41°C (1.31°C–1.51°C), where the range represents**
16 **the spread in the datasets' median estimates.** Analyses of paleo records, historical observations, model
17 simulations, and underlying physical principles are all in agreement that LSATs are increasing at a higher
18 rate than SST as a result of differences in evaporation, land-climate feedbacks, and changes in the aerosol
19 forcing over land (*very high confidence*). For the 2000–2016 period, the land-to-ocean warming ratio (about
20 1.6) is in close agreement between different observational records and the CMIP5 climate model simulations
21 (the *likely* range of 1.54 to 1.81). {2.2.1}

22
23 **Anthropogenic warming has resulted in shifts of climate zones, primarily as an increase in dry**
24 **climates and decrease of polar climates (*high confidence*). Ongoing warming is projected to result in**
25 **new, hot climates in tropical regions and to shift climate zones poleward in the mid- to high latitudes**
26 **and upward in regions of higher elevation (*high confidence*).** Ecosystems in these regions will become
27 increasingly exposed to temperature and rainfall extremes beyond climate regimes they are currently adapted
28 to (*high confidence*), which can alter their structure, composition and functioning. Additionally, high-latitude
29 warming is projected to accelerate permafrost thawing and increase disturbance in boreal forests through
30 abiotic (e.g., drought, fire) and biotic (e.g., pests, disease) agents (*high confidence*). {2.2.1, 2.2.2, 2.5.3}

31
32 **Globally, greening trends (trends of increased photosynthetic activity in vegetation) have increased**
33 **over the last 2-3 decades by 22–33%, particularly over China, India, many parts of Europe, central**
34 **North America, southeast Brazil and southeast Australia (*high confidence*).** This results from a
35 combination of direct (i.e., land use and management, forest conservation and expansion) and indirect factors
36 (i.e., CO₂ fertilisation, extended growing season, global warming, nitrogen deposition, increase of diffuse
37 radiation) linked to human activities (*high confidence*). Browning trends (trends of decreasing photosynthetic
38 activity) are projected in many regions where increases in drought and heat waves are projected in a warmer
39 climate. There is *low confidence* in the projections of global greening and browning trends. {2.2.4, Cross-
40 Chapter Box 4: Climate change and urbanisation, in this chapter}

41
42 **The frequency and intensity of some extreme weather and climate events have increased as a**
43 **consequence of global warming and will continue to increase under medium and high emission**
44 **scenarios (*high confidence*).** Recent heat-related events, e.g., heat waves, have been made more frequent or
45 intense due to anthropogenic greenhouse gas emissions in most land regions and the frequency and intensity
46 of drought has increased in Amazonia, north-eastern Brazil, the Mediterranean, Patagonia, most of Africa
47 and north-eastern China (*medium confidence*). Heat waves are projected to increase in frequency, intensity
48 and duration in most parts of the world (*high confidence*) and drought frequency and intensity is projected to
49 increase in some regions that are already drought prone, predominantly in the Mediterranean, central Europe,
50 the southern Amazon and southern Africa (*medium confidence*). These changes will impact ecosystems, food
51 security and land processes including greenhouse gas (GHG) fluxes (*high confidence*). {2.2.5}

1
2 **Climate change is playing an increasing role in determining wildfire regimes along-side human**
3 **activity (*medium confidence*), with future climate variability expected to enhance the risk and severity**
4 **of wildfires in many biomes such as tropical rainforests (*high confidence*).** Fire weather seasons have
5 lengthened globally between 1979 and 2013 (*low confidence*). Global land area burned has declined in recent
6 decades, mainly due to less burning in grasslands and savannahs (*high confidence*). While drought remains
7 the dominant driver of fire emissions, there has recently been increased fire activity in some tropical and
8 temperate regions during normal to wetter than average years due to warmer temperatures that increase
9 vegetation flammability (*medium confidence*). The boreal zone is also experiencing larger and more frequent
10 fires, and this may increase under a warmer climate (*medium confidence*). {Cross-Chapter Box 4: Climate
11 change and urbanisation, in this chapter}

12 ***Terrestrial greenhouse gas fluxes on unmanaged and managed lands***

13 **Agriculture, Forestry and Other Land Use (AFOLU) is a significant net source of GHG emissions**
14 **(*high confidence*), contributing to about 22% of anthropogenic emissions of carbon dioxide (CO₂),**
15 **methane (CH₄), and nitrous oxide (N₂O) combined as CO₂ equivalents in 2007 to 2016 (*medium***
16 ***confidence*).** AFOLU results in both emissions and removals of CO₂, CH₄, and N₂O to and from the
17 atmosphere (*high confidence*). These fluxes are affected simultaneously by natural and human drivers,
18 making it difficult to separate natural from anthropogenic fluxes (*very high confidence*). {2.3}

19 **The total net land-atmosphere flux of CO₂ on both managed and unmanaged lands very likely**
20 **provided a global net removal from 2008 to 2017 according to models, (-6.2 ± 3.7 GtCO₂ yr⁻¹, *medium***
21 ***confidence*).** This net removal is comprised of two major components: i) modelled net anthropogenic
22 emissions from AFOLU are *likely* 5.5 ± 2.6 GtCO₂ yr⁻¹ driven by land cover change, including deforestation
23 and afforestation/reforestation, and wood harvesting (accounting for about 13% of total net anthropogenic
24 emissions of CO₂) (*medium confidence*); and ii) modelled net removals due to non-anthropogenic processes
25 are *likely* 11.7 ± 2.6 GtCO₂ yr⁻¹ on managed and unmanaged lands, driven by environmental changes such as
26 increasing CO₂, nitrogen deposition, and changes in climate (accounting for a removal of 29% of the CO₂
27 emitted from all anthropogenic activities (fossil fuel, industry and AFOLU) (*medium confidence*). {2.3.1}

28 **The anthropogenic emissions of CO₂ from AFOLU reported in countries' GHG inventories were 0.1 ±**
29 **1.0 GtCO₂ yr⁻¹ globally during 2005 to 2014 (*low confidence*), much lower than emission estimates**
30 **from global models of 5.1 ± 2.6 GtCO₂ yr⁻¹ over the same time period. Reconciling these differences**
31 **can support consistency and transparency in assessing global progress towards meeting modelled**
32 **mitigation pathway such as under the Paris Agreement's global stocktake (*medium confidence*).** This
33 discrepancy is consistent with understanding of the different approaches used to defining anthropogenic
34 fluxes. Inventories consider larger areas of forested lands as managed than models do, and report all fluxes
35 on managed lands as anthropogenic, including a large net sink due to the indirect effects of changing
36 environmental conditions (e.g., climate change, and change in atmospheric CO₂ and N). In contrast, the
37 models assign part of this indirect forest sink to the non-anthropogenic sink on unmanaged lands. {2.3.1}

38 **The gross emissions from AFOLU (one third of total global emissions) are more indicative of**
39 **mitigation potential of reduced deforestation than the global net emissions (13% of total global**
40 **emissions), which include compensating deforestation and afforestation fluxes (*high confidence*).** The
41 net flux of CO₂ from AFOLU is composed of two opposing gross fluxes: gross emissions (20 GtCO₂ yr⁻¹)
42 from deforestation, cultivation of soils, and oxidation of wood products; and gross removals (-14 GtCO₂ yr⁻¹)
43 largely from forest growth following wood harvest and agricultural abandonment (*medium confidence*).
44 {2.3.1}

45 **Land is a net source of CH₄, accounting for 61% of anthropogenic CH₄ emissions for the 2005–2015**

1 **period (*medium confidence*)**. The pause in the rise of atmospheric CH₄ concentrations between 2000 and
2 2006 and the subsequent renewed increase appear to be partially associated with land use and land use
3 change. The recent depletion trend of the ¹³C isotope in the atmosphere indicates that higher biogenic sources
4 explain part of the current CH₄ increase and that biogenic sources make up a larger proportion of the source
5 mix than they did before 2000 (*high confidence*). In agreement with the findings of AR5, tropical wetlands
6 and peatlands continue to be important drivers of inter-annual variability and current CH₄ concentration
7 increases (*medium evidence, high agreement*). Ruminants and the expansion of rice cultivation are also
8 important contributors to the current trend (*medium evidence, high agreement*). There is significant and
9 ongoing accumulation of CH₄ in the atmosphere (*very high confidence*). {2.3.2}

10
11 **AFOLU is the main anthropogenic sources of N₂O primarily due to nitrogen (N) application to soils**
12 (*high confidence*). In croplands, the main driver of N₂O emissions is a lack of synchronisation between crop
13 N demand and soil N supply, with approximately 50% of the N applied to agricultural land not taken up by
14 the crop. Cropland soils emit over 3 Mt N₂O-N yr⁻¹ (*medium confidence*). Because the response of N₂O
15 emissions to fertiliser application rates is non-linear, in regions of the World where low N application rates
16 dominate, such as sub-Saharan Africa and parts of Eastern Europe, increases in N fertiliser use would
17 generate relatively small increases in agricultural N₂O emissions. Decreases in application rates in regions
18 where application rates are high and exceed crop demand for parts of the growing season will have very
19 large effects on emissions reductions (*medium evidence, high agreement*). {2.3.3}

20
21 **While managed pastures make up only one-quarter of grazing lands, they contributed more than**
22 **three-quarters of N₂O emissions from grazing lands between 1961 and 2014 with rapid recent**
23 **increases of N inputs resulting in disproportionate growth in emissions from these lands (*medium***
24 ***confidence*)**. Grazing lands (pastures and rangelands) are responsible for more than one-third of total
25 anthropogenic N₂O emissions or more than one-half of agricultural emissions (*high confidence*). Emissions
26 are largely from North America, Europe, East Asia, and South Asia, but hotspots are shifting from Europe to
27 southern Asia (*medium confidence*). {2.3.3}

28
29 **Increased emissions from vegetation and soils due to climate change in the future are expected to**
30 **counteract potential sinks due to CO₂ fertilisation (*low confidence*)**. Responses of vegetation and soil
31 organic carbon (SOC) to rising atmospheric CO₂ concentration and climate change are not well constrained
32 by observations (*medium confidence*). Nutrient (e.g., nitrogen, phosphorus) availability can limit future plant
33 growth and carbon storage under rising CO₂ (*high confidence*). However, new evidence suggests that
34 ecosystem adaptation through plant-microbe symbioses could alleviate some nitrogen limitation (*medium*
35 *evidence, high agreement*). Warming of soils and increased litter inputs will accelerate carbon losses through
36 microbial respiration (*high confidence*). Thawing of high-latitude/altitude permafrost will increase rates of
37 SOC loss and change the balance between CO₂ and CH₄ emissions (*medium confidence*). The balance
38 between increased respiration in warmer climates and carbon uptake from enhanced plant growth is a key
39 uncertainty for the size of the future land carbon sink (*medium confidence*). {2.3.1, 2.7.2, Box 2.3}

40 41 ***Biophysical and biogeochemical land forcing and feedbacks to the climate system***

42
43 **Changes in land conditions from human use or climate change in turn affect regional and global**
44 **climate (*high confidence*)**. On the global scale, this is driven by changes in emissions or removals of CO₂,
45 CH₄, and N₂O by land (biogeochemical effects) and by changes in the surface albedo (*very high confidence*).
46 Any local land changes that redistribute energy and water vapour between the land and the atmosphere
47 influence regional climate (biophysical effects; *high confidence*). However, there is *no confidence* in whether
48 such biophysical effects influence global climate. {2.1, 2.3, 2.5.1, 2.5.2}

49
50 **Changes in land conditions modulate the likelihood, intensity and duration of many extreme events**
51 **including heat waves (*high confidence*) and heavy precipitation events (*medium confidence*)**. Dry soil

1 conditions favour or strengthen summer heat wave conditions through reduced evapotranspiration and
2 increased sensible heat. By contrast wet soil conditions, for example from irrigation, or crop management
3 practices that maintain a cover crop all year round, can dampen extreme warm events through increased
4 evapotranspiration and reduced sensible heat. Droughts can be intensified by poor land management.
5 Urbanisation increases extreme rainfall events over or downwind of cities (*medium confidence*). {2.5.1,
6 2.5.2, 2.5.3}

7
8 **Historical changes in anthropogenic land cover have resulted in a mean annual global warming of**
9 **surface air from biogeochemical effects (*very high confidence*), dampened by a cooling from**
10 **biophysical effects (*medium confidence*).** Biogeochemical warming results from increased emissions of
11 GHGs by land, with model-based estimates of $+0.20\pm 0.05^{\circ}\text{C}$ (global climate models) and $+0.24\pm 0.12^{\circ}\text{C}$
12 (dynamic global vegetation models, DGVMs) as well as an observation-based estimate of $+0.25\pm 0.10^{\circ}\text{C}$. A
13 net biophysical cooling of $-0.10\pm 0.14^{\circ}\text{C}$ has been derived from global climate models in response to the
14 increased surface albedo and decreased turbulent heat fluxes, but it is smaller than the warming effect from
15 land-based emissions. However when both biogeochemical and biophysical effects are accounted for within
16 the same global climate model, the models do not agree on the sign of the net change in mean annual surface
17 air temperature. {2.3, 2.5.1, Box 2.1}

18
19 **The future projected changes in anthropogenic land cover that have been examined for AR5 would**
20 **result in a biogeochemical warming and a biophysical cooling whose magnitudes depend on the**
21 **scenario (*high confidence*).** Biogeochemical warming has been projected for RCP8.5 by both global climate
22 models ($+0.20\pm 0.15^{\circ}\text{C}$) and DGVMs ($+0.28\pm 0.11^{\circ}\text{C}$) (*high confidence*). A global biophysical cooling of
23 $0.10\pm 0.14^{\circ}\text{C}$ is estimated from global climate models, and projected to dampen the land-based warming (*low*
24 *confidence*). For RCP4.5 the biogeochemical warming estimated from global climate models ($+0.12\pm 0.17^{\circ}\text{C}$)
25 is stronger than the warming estimated by DGVMs ($+0.01\pm 0.04^{\circ}\text{C}$) but based on *limited evidence*, as is the
26 biophysical cooling ($-0.10\pm 0.21^{\circ}\text{C}$). {2.5.2}

27
28 **Regional climate change can be dampened or enhanced by changes in local land cover and land use**
29 **(*high confidence*) but this depends on the location and the season (*high confidence*).** In boreal regions,
30 for example, where projected climate change will migrate treeline northward, increase the growing season
31 length and thaw permafrost, regional winter warming will be enhanced by decreased surface albedo and
32 snow, whereas warming will be dampened during the growing season due to larger evapotranspiration (*high*
33 *confidence*). In the tropics, wherever climate change will increase rainfall, vegetation growth and associated
34 increase in evapotranspiration will result in a dampening effect on regional warming (*medium confidence*).
35 {2.5.2, 2.5.3}

36
37 **According to model-based studies, changes in local land cover or available water from irrigation**
38 **affect climate in regions as far as few hundreds of kilometres downwind (*high confidence*).** The local
39 redistribution of water and energy following the changes on land affect the horizontal and vertical gradients
40 of temperature, pressure and moisture, thus alter regional winds and consequently moisture and temperature
41 advection and convection, and this affects precipitation. {2.5.2, 2.5.4, Cross-Chapter Box 4: Climate
42 Change and Urbanisation}

43
44 **Future increases in both climate change and urbanisation will enhance warming in cities and their**
45 **surroundings (urban heat island), especially during heat waves (*high confidence*).** Urban and peri-urban
46 agriculture, and more generally urban greening, can contribute to mitigation (*medium confidence*) as well as
47 to adaptation (*high confidence*), with co-benefits for food security and reduced soil-water-air pollution.
48 {Cross-Chapter Box 4: Climate Change and Urbanisation}

49
50 **Regional climate is strongly affected by natural land aerosols (*medium confidence*) (e.g., mineral dust,**
51 **black, brown and organic carbon), but there is *low confidence* in historical trends, interannual and**

1 **decadal variability, and future changes.** Forest cover affects climate through emissions of biogenic
2 volatile organic compounds (BVOC) and aerosols (*low confidence*). The decrease in the emissions of
3 BVOC resulting from the historical conversion of forests to cropland has resulted in a positive radiative
4 forcing through direct and indirect aerosol effects, a negative radiative forcing through the reduction in the
5 atmospheric lifetime of methane and it has contributed to increased ozone concentrations in different
6 regions (*low confidence*). {2.4, 2.5}

7
8 ***Consequences for the climate system of land-based adaptation and mitigation options, including carbon
9 dioxide removal (negative emissions)***

10
11 **About one quarter of the 2030 mitigation pledged by countries in their initial Nationally Determined
12 Contributions (NDCs) under the Paris Agreement is expected to come from land-based mitigation
13 options (*medium confidence*).** Most of the Nationally Determined Contributions (NDCs) submitted by
14 countries include land-based mitigation, although many lack details. Several refer explicitly to reduced
15 deforestation and forest sinks, while a few include soil carbon sequestration, agricultural management and
16 bioenergy. Full implementation of NDCs (submitted by February 2016) is expected to result in net
17 removals of 0.4–1.3 GtCO₂ y⁻¹ in 2030 compared to the net flux in 2010, where the range represents low to
18 high mitigation ambition in pledges, not uncertainty in estimates (*medium confidence*). {2.6.3}

19
20 **Several mitigation response options have technical potential for >3 GtCO₂-eq yr⁻¹ by 2050 through
21 reduced emissions and Carbon Dioxide Removal (CDR) (*high confidence*), some of which compete
22 for land and other resources, while others may reduce the demand for land (*high confidence*).**
23 Estimates of the technical potential of individual response options are not necessarily additive. The largest
24 potential for reducing AFOLU emissions are through reduced deforestation and forest degradation (0.4–5.8
25 GtCO₂-eq yr⁻¹) (*high confidence*), a shift towards plant-based diets (0.7–8.0 GtCO₂-eq yr⁻¹) (*high
26 confidence*) and reduced food and agricultural waste (0.8–4.5 CO₂-eq yr⁻¹) (*high confidence*). Agriculture
27 measures combined could mitigate 0.3–3.4 GtCO₂-eq yr⁻¹ (*medium confidence*). The options with largest
28 potential for CDR are afforestation/reforestation (0.5–10.1 CO₂-eq yr⁻¹) (*medium confidence*), soil carbon
29 sequestration in croplands and grasslands (0.4–8.6 CO₂-eq yr⁻¹) (*high confidence*) and Bioenergy with
30 Carbon Capture and Storage (BECCS) (0.4–11.3 CO₂-eq yr⁻¹) (*medium confidence*). While some estimates
31 include sustainability and cost considerations, most do not include socio-economic barriers, the impacts of
32 future climate change or non-GHG climate forcings. {2.6.1}

33
34 **Response options intended to mitigate global warming will also affect the climate locally and
35 regionally through biophysical effects (*high confidence*).** Expansion of forest area, for example, typically
36 removes CO₂ from the atmosphere and thus dampens global warming (biogeochemical effect, *high
37 confidence*), but the biophysical effects can dampen or enhance regional warming depending on location,
38 season and time of day. During the growing season, afforestation generally brings cooler days from
39 increased evapotranspiration, and warmer nights (*high confidence*). During the dormant season, forests are
40 warmer than any other land cover, especially in snow-covered areas where forest cover reduces albedo
41 (*high confidence*). At the global level, the temperature effects of boreal afforestation/reforestation run
42 counter to GHG effects, while in the tropics they enhance GHG effects. In addition, trees locally dampen
43 the amplitude of heat extremes (*medium confidence*). {2.5.2, 2.5.4, 2.7, Cross-Chapter Box 4: Climate
44 Change and Urbanisation}

45
46 **Mitigation response options related to land use are a key element of most modelled scenarios that
47 provide strong mitigation, alongside emissions reduction in other sectors (*high confidence*).** More
48 stringent climate targets rely more heavily on land-based mitigation options, in particular, CDR
49 (*high confidence*). Across a range of scenarios in 2100, CDR is delivered by both afforestation (median
50 values of -1.3, -1.7 and -2.4 GtCO₂yr⁻¹ for scenarios RCP4.5, RCP2.6 and RCP1.9 respectively) and
51 bioenergy with carbon capture and storage (BECCS) (-6.5, -11 and -14.9 GtCO₂ yr⁻¹). Emissions of CH₄

1 and N₂O are reduced through improved agricultural and livestock management as well as dietary shifts
2 away from emission-intensive livestock products by 133.2, 108.4 and 73.5 MtCH₄yr⁻¹; and 7.4, 6.1 and 4.5
3 MtN₂O yr⁻¹ for the same set of scenarios in 2100 (*high confidence*). High levels of bioenergy crop
4 production can result in increased N₂O emissions due to fertiliser use. The Integrated Assessment Models
5 that produce these scenarios mostly neglect the biophysical effects of land-use on global and regional
6 warming. {2.5, 2.6.2}

7
8 **Large-scale implementation of mitigation response options that limit warming to 1.5 or 2°C would**
9 **require conversion of large areas of land for afforestation/reforestation and bioenergy crops, which**
10 **could lead to short-term carbon losses (*high confidence*).** The change of global forest area in mitigation
11 pathways ranges from about -0.2 to +7.2 Mkm² between 2010 and 2100 (median values across a range of
12 models and scenarios: RCP4.5, RCP2.6, RCP1.9), and the land demand for bioenergy crops ranges from
13 about 3.2–6.6 Mkm² in 2100 (*high confidence*). Large-scale land-based CDR is associated with multiple
14 feasibility and sustainability constraints (Chapters 6, 7). In high carbon lands such as forests and peatlands,
15 the carbon benefits of land protection are greater in the short-term than converting land to bioenergy crops
16 for BECCS, which can take several harvest cycles to ‘pay-back’ the carbon emitted during conversion
17 (carbon-debt), from decades to over a century (*medium confidence*). {2.6.2, Chapters 6, 7}

18
19 **It is possible to achieve climate change targets with low need for land-demanding CDR such as**
20 **BECCS, but such scenarios rely more on rapidly reduced emissions or CDR from forests, agriculture**
21 **and other sectors.** Terrestrial CDR has the technical potential to balance emissions that are difficult to
22 eliminate with current technologies (including food production). Scenarios that achieve climate change
23 targets with less need for terrestrial CDR rely on agricultural demand-side changes (diet change, waste
24 reduction), and changes in agricultural production such as agricultural intensification. Such pathways that
25 minimise land use for bioenergy and BECCS are characterised by rapid and early reduction of GHG
26 emissions in all sectors, as well as earlier CDR in through afforestation. In contrast, delayed mitigation
27 action would increase reliance on land-based CDR (*high confidence*). {2.6.2}

2.1 Introduction: Land – climate interactions

This chapter assesses the literature on two-way interactions between climate and land, with focus on scientific findings published since AR5 and some aspects of the land-climate interactions that were not assessed in previous IPCC reports. Previous IPCC assessments recognised that climate affects land cover and land surface processes, which in turn affect climate. However, previous assessments mostly focused on the contribution of land to global climate change via its role in emitting and absorbing greenhouse gases (GHGs) and short-lived climate forcers (SLCFs), or via implications of changes in surface reflective properties (i.e., albedo) for solar radiation absorbed by the surface. This chapter examines scientific advances in understanding the interactive changes of climate and land, including impacts of climate change, variability and extremes on managed and unmanaged lands. It assesses climate forcing of land changes from direct (e.g., land use change and land management) and indirect (e.g., increasing atmospheric CO₂ concentration and nitrogen deposition) effects at local, regional, and global scale.

2.1.1 Recap of previous IPCC and other relevant reports as baselines

The evidence that land cover matters for the climate system have long been known, especially from early paleoclimate modelling studies and impacts of human-induced deforestation at the margin of deserts (de Noblet et al. 1996; Kageyama et al. 2004). The understanding of how land use activities impact climate has been put forward by the pioneering work of (Charney 1975) who examined the role of overgrazing-induced desertification on the Sahelian climate.

Since then there have been many modelling studies that reported impacts of idealised or simplified land cover changes on weather patterns (e.g., Pielke et al. 2011). The number of studies dealing with such issues has increased significantly over the past 10 years, with more studies that address realistic past or projected land changes. However, very few studies have addressed the impacts of land cover changes on climate as very few land surface models embedded within climate models (whether global or regional), include a representation of land management. Observation-based evidence of land-induced climate impacts emerged even more recently (e.g., Alkama and Cescatti 2016; Bright et al. 2017; Lee et al. 2011; Li et al. 2015; Duveiller et al. 2018; Forzieri et al. 2017) and the literature is therefore limited.

In previous IPCC reports, the interactions between climate change and land were covered separately by three working groups. AR5 WGI assessed the role of land use change in radiative forcing, land-based GHGs source and sink, and water cycle changes that focused on changes of evapotranspiration, snow and ice, runoff, and humidity. AR5 WGII examined impacts of climate change on land, including terrestrial and freshwater ecosystems, managed ecosystems, and cities and settlements. AR5 WGIII assessed land-based climate change mitigation goals and pathways in the AFOLU. Here, this chapter assess land-climate interactions from all three working groups. It also builds on previous special reports such as the Special Report on Global Warming of 1.5°C (SR15). It links to the IPCC Guidelines on National Greenhouse Gas Inventories in the land sector. Importantly, this chapter assesses knowledge that has never been reported in any of those previous reports. Finally, the chapter also tries to reconcile the possible inconsistencies across the various IPCC reports.

Land-based water cycle changes: AR5 reported an increase in global evapotranspiration from the early 1980s to 2000s, but a constraint on further increases from low of soil moisture availability. Rising CO₂ concentration limits stomatal opening and thus also reduces transpiration, a component of evapotranspiration. Increasing aerosol levels, and declining surface wind speeds and levels of solar radiation reaching the ground are additional regional causes of the decrease in evapotranspiration.

Land area precipitation change: Averaged over the mid-latitude land areas of the Northern Hemisphere, precipitation has increased since 1901 (*medium confidence* before and *high confidence* after 1951). For other

1 latitudes, area-averaged long-term positive or negative trends have *low confidence*. There are *likely* more
2 land regions where the number of heavy precipitation events has increased than where it has decreased.
3 Extreme precipitation events over most of the mid-latitude land masses and over wet tropical regions will
4 very likely become more intense and more frequent (IPCC 2013a).

5
6 **Land-based GHGs:** AR5 reported that annual net CO₂ emissions from anthropogenic land use change were
7 0.9 [0.1–1.7] GtC yr⁻¹ on average during 2002 to 2011 (*medium confidence*). From 1750 to 2011, CO₂
8 emissions from fossil fuel combustion have released an estimated 375 [345–405] GtC to the atmosphere,
9 while deforestation and other land use change have released an estimated 180 [100–260] GtC. Of these
10 cumulative anthropogenic CO₂ emissions, 240 [230–250] GtC have accumulated in the atmosphere, 155
11 [125–185] GtC have been taken up by the ocean and 160 [70–250] GtC have accumulated in terrestrial
12 ecosystems (i.e., the cumulative residual land sink) (Ciais et al. 2013a). Updated assessment and knowledge
13 gaps are covered in Section 2.3.

14
15 **Future terrestrial carbon source/sink:** AR5 projected with *high confidence* that tropical ecosystems will
16 uptake less carbon and with *medium confidence* that at high latitudes, land carbon sink will increase in a
17 warmer climate. Thawing permafrost in the high latitudes is potentially a large carbon source at warmer
18 climate, but the magnitude of CO₂ and CH₄ emissions due to permafrost thawing is still uncertain. The SR15
19 further indicates that constraining warming to 1.5°C would prevent the melting of an estimated permafrost
20 area of 2 million km² over the next centuries compared to 2°C. Updates to these assessments are found in
21 Sections 2.3.

22
23 **Land use change altered albedo:** AR5 stated with *high confidence* that anthropogenic land use change has
24 increased the land surface albedo, which has led to a RF of $-0.15 \pm 0.10 \text{ W m}^{-2}$. However, it also underlined
25 that the sources of the large spread across independent estimates was caused by differences in assumptions
26 for the albedo of natural and managed surfaces and for the fraction of land use change before 1750.
27 Generally, our understanding of albedo changes from land use change has been enhanced from AR4 to AR5,
28 with a narrower range of estimates and a higher confidence level. The radiative forcing from changes in
29 albedo induced by land use changes was estimated in AR5 at -0.15 W m^{-2} (-0.25 to about -0.05), with
30 *medium confidence* in AR5 (Shindell et al. 2013). This was an improvement over AR4 in which it was
31 estimated at -0.2 W m^{-2} (-0.4 to about 0), with *low to medium confidence* (Forster et al. 2007). Section 2.5
32 shows that albedo is not the only source of biophysical land-based climate forcing to be considered.

33
34 **Hydrological feedback to climate:** Land use changes also affect surface temperatures through non-radiative
35 processes, and particularly through the hydrological cycle. These processes are less well known and are
36 difficult to quantify, but tend to offset the impact of albedo changes. As a consequence, there is low
37 agreement on the sign of the net change in global mean temperature as a result of land use change (Hartmann
38 et al. 2013a). An updated assessment on these points is covered in Section 2.5 and 2.2

39
40 **Climate-related extremes on land:** AR5 reported that impacts from recent climate-related extremes reveal
41 significant vulnerability and exposure of some ecosystems to current climate variability. Impacts of such
42 climate-related extremes include alteration of ecosystems, disruption of food production and water supply,
43 damage to infrastructure and settlements, morbidity and mortality, and consequences for mental health and
44 human well-being (Burkett et al. 2014). The SR15 further indicates that limiting global warming to 1.5°C
45 limits the risks of increases in heavy precipitation events in several regions (*high confidence*). In urban areas
46 climate change is projected to increase risks for people, assets, economies and ecosystems (*very high*
47 *confidence*). These risks are amplified for those lacking essential infrastructure and services or living in
48 exposed areas. Updated assessment and knowledge gap for this chapter are covered in Section 2.2 and Cross-
49 Chapter Box 4: Climate Change and Urbanisation.

50
51 **Land-based climate change adaptation and mitigation:** AR5 reported that adaptation and mitigation

1 choices in the near-term will affect the risks related to climate change throughout the 21st century (Burkett et
2 al. 2014). Agriculture, forestry and other land use (AFOLU) are responsible for about 10–12 GtCO₂eq yr⁻¹
3 anthropogenic greenhouse gas emissions mainly from deforestation and agricultural production. Global CO₂
4 emissions from forestry and other land use have declined since AR4, largely due to increased afforestation.
5 The SR15 further indicates that afforestation and bioenergy with carbon capture and storage (BECCS) are
6 important land-based carbon dioxide removal (CDR) options. It also states that land use and land-use change
7 emerge as a critical feature of virtually all mitigation pathways that seek to limit global warming to 1.5°C.
8 Climate Change 2014 Synthesis Report concluded that co-benefits and adverse side effects of mitigation
9 could affect achievement of other objectives such as those related to human health, food security,
10 biodiversity, local environmental quality, energy access, livelihoods and equitable sustainable development.
11 Updated assessment and knowledge gaps are covered in Section 2.6 and Chapter 7.

12
13 Overall, sustainable land management is largely constrained by climate change and extremes, but also puts
14 bounds on the capacity of land to effectively adapt to climate change and mitigate its impacts. Scientific
15 knowledge has advanced on how to optimise our adaptation and mitigation efforts while coordinating
16 sustainable land management across sectors and stakeholder. Details are assessed in subsequent sections.

17 **2.1.2 Introduction to the chapter structure**

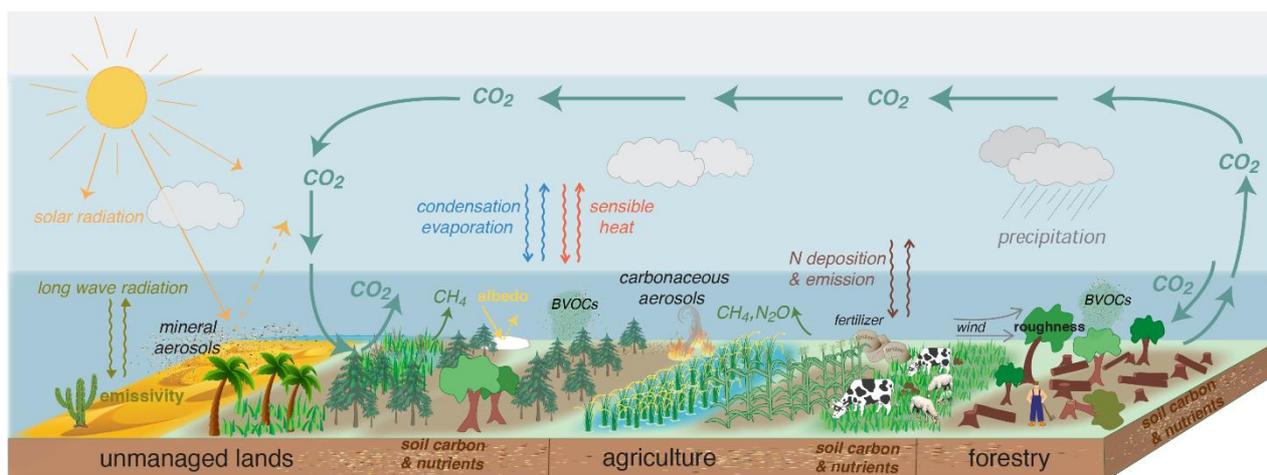
18
19
20 This chapter assess the consequences of changes in land cover and functioning, resulting from both land use
21 and climate change, to global and regional climates. The chapter starts by an assessment of the historical and
22 projected responses of land-based and processes to climate change and extremes (Section 2.2). Subsequently,
23 the chapter assesses historical and future changes in terrestrial GHG fluxes (Section 2.3), non-GHG fluxes
24 and precursors of SLCFs (Section 2.4. Section 2.4 focuses on how historical and future changes in land use
25 and land cover influence climate change/variability through biophysical and biogeochemical forcing and
26 feedbacks, how specific land management affects climate, and how in turn climate-induced land changes
27 feedback to climate. Section 2.6 assesses consequences of land-based adaptation and mitigation options for
28 the climate system in GHG and non-GHG exchanges. Sections 2.3 and 2.6 addresses implications of the
29 Paris Agreement for land-climate interactions, and the scientific evidence base for ongoing negotiations
30 around the Paris rulebook, the Global Stocktake, and credibility in measuring, reporting and verifying the
31 climate impacts of anthropogenic activities on land. The chapter also examines how land use and
32 management practices may affect climate change through biophysical feedbacks and radiative forcing
33 (Section 2.5), and assesses policy relevant projected land use changes and sustainable land management for
34 mitigation and adaptation (Section 2.6). Finally, the chapter concludes with a brief assessment of advances in
35 the understanding of ecological and biogeochemical processes underlying land-climate interactions (Section
36 2.7).

37
38 The chapter includes three chapter boxes providing general overview of (i) processes underlying land-
39 climate interactions (Box 2.1.); (ii) methodological approaches for estimating anthropogenic land carbon
40 fluxes from national to global scales (Box 2.2.); (iii) CO₂ fertilisation and enhanced terrestrial uptake of
41 carbon (Box 2.3). In addition this chapter includes two cross-chapter boxes on climate change and fire
42 (Cross-Chapter Box 3); and on urbanisation and climate change (Cross-Chapter Box 4).

43
44 In summary, the chapter assesses scientific understanding related to: 1) how a changing climate affects
45 terrestrial ecosystems, including those on managed lands; 2) how land affects climate through biophysical
46 and biogeochemical feedbacks; and 3) how land use or cover change and land management play an
47 important and complex role in the climate system. This chapter also pays special attention to advances in
48 understanding cross-scale interactions, emerging issues, heterogeneity, and teleconnections.

Box 2.1: Processes underlying land-climate interactions

Land continuously interacts with the atmosphere through exchanges of, for instance, greenhouse gases (e.g., CO₂, CH₄, N₂O), water, energy, or precursors of short lived-climate forcers (e.g., biogenic volatile organic compounds, dust, black carbon). The terrestrial biosphere also interacts with oceans through processes such as the influx of freshwater, nutrients, carbon and particles. These interactions affect where and when rain falls and thus irrigation needs for crops, frequency and intensity of heat waves, and air quality. They are modified by global and regional climate change, decadal, interannual and seasonal climatic variations, and weather extremes, as well as human actions on land (e.g., crop and forest management, afforestation and deforestation). This in turn affects atmospheric composition, surface temperature, hydrological cycle and thus local, regional and global climate. This box introduces some of the fundamental land processes governing biophysical and biogeochemical effects and feedbacks to the climate (Box 2.1 Figure 1)



Box 2.1 Figure 1 The structure and functioning of managed and unmanaged ecosystems that affect local, regional, and global climate. Land surface characteristics such as albedo and emissivity determine the amount of solar and long-wave radiation absorbed by land and reflected or emitted to the atmosphere. Surface roughness influences turbulent exchanges of momentum, energy, water, and biogeochemical tracers. Land ecosystems modulate the atmospheric composition through emissions and removals of many GHGs and precursors of SLCFs, including biogenic volatile organic compounds (BVOCs) and mineral dust. Atmospheric aerosols formed from these precursors affect regional climate by altering amounts of precipitation and of radiation reaching land surfaces through their role in clouds physics.

‘Biophysical interactions’ are exchanges of water and energy between the land and the atmosphere (Section 2.5). Land warms up from absorbing solar and long-wave radiation; it cools down through transfers of sensible heat (*via* conduction and convection) and latent heat (energy associated with water evapotranspiration) to the atmosphere and through longwave radiation emission from the land surface (Box 2.1 Figure 1). These interactions between the land and the atmosphere depend on the land surface characteristics, including reflectivity of short-wave radiation (albedo), emissivity of long wave radiation by vegetation and soils, surface roughness, and soil water access by vegetation, which depends on both soil characteristics and amounts of roots. Over seasonal, interannual and decadal time scales, these characteristics vary among different land cover and land-use types and are affected by both natural processes and land management (Anderson et al. 2011). A dense vegetation with high leaf area index, like forests, may absorb more energy than nearby herbaceous vegetation partly due to differences in surface albedo (especially when snow is on the ground). However, denser vegetation also sends more energy back to the atmosphere in the form of evapotranspiration (Bonan, 2008; Burakowski et al., 2018; Ellison et al., 2017; Section 2.5.2) and this contributes to changes in atmospheric water vapour content, affecting and rainfall.

Particularly in extra-tropical regions, these characteristics exhibit strong seasonal patterns with the development and senescence of the vegetation (e.g., leaf colour change and drop). For example, in deciduous forests, seasonal growth increases albedo by 20–50% from the spring minima to growing season maxima, followed by rapid decrease during leaf fall, whereas in grasslands, spring greening causes albedo decreases and only increases with vegetation browning (Hollinger et al. 2010). The seasonal patterns of sensible and

1 latent heat fluxes are also driven by the cycle of leaf development and senescence in temperate deciduous
2 forests: sensible heat fluxes peak in spring and autumn and latent heat fluxes peak in mid-summer (Moore et
3 al. 1996; Richardson et al. 2013).

4
5 Exchanges of greenhouse gases between the land and the atmosphere are referred to as ‘biogeochemical
6 interactions’ (Section 2.3), which are driven mainly by the balance between photosynthesis and respiration
7 by plants, and by the decomposition of soil organic matter by microbes. The conversion of atmospheric
8 carbon dioxide into organic compounds by plant photosynthesis, known as terrestrial net primary
9 productivity, is the source of plant growth, food for human and other organisms, and soil organic carbon.
10 Due to strong seasonal patterns of growth, northern hemisphere terrestrial ecosystems are largely responsible
11 for the seasonal variations in global atmospheric CO₂ concentrations. In addition to CO₂, soils emit
12 methane (CH₄) and nitrous oxide (N₂O) (Section 2.3). Soil temperature and moisture strongly affect
13 microbial activities and resulting fluxes of these three greenhouse gases.

14
15 Much like fossil fuel emissions, GHG emissions from anthropogenic land cover change and land
16 management are ‘forcers’ on the climate system. Other land-based changes to climate are described as
17 ‘feedbacks’ to the climate system - a process by which climate change influences some property of land,
18 which in turn diminishes (negative feedback) or amplifies (positive feedback) climate change. Examples of
19 feedbacks include the changes in the strength of land carbon sinks or sources, soil moisture and plant
20 phenology (Section 2.5.3).

21
22 Incorporating these land-climate processes into climate projections allows for increased understanding of the
23 land’s response to climate change (Section 2.2), and to better quantify the potential of land-based response
24 options for climate change mitigation (Section 2.6). However, to date Earth system models (ESMs)
25 incorporate some combined biophysical and biogeochemical processes only to limited extent and many
26 relevant processes about how plants and soils interactively respond to climate changes are still to be
27 included. (Section 2.7). And even within this class of models, the spread in ESM projections is large, in part
28 because of their varying ability to represent land-climate processes (Hoffman et al. 2014). Significant
29 progress in understanding of these processes has nevertheless been made since AR5.

2.2 The effect of climate variability and change on land

2.2.1 Overview of climate impacts on land

2.2.1.1 Climate drivers of land form and function

Energy is redistributed from the warm equator to the colder poles through large-scale atmospheric and oceanic processes driving the Earth's weather and climate (Oort and Peixóto 1983; Carissimo et al. 1985; Yang et al. 2015a). Subsequently, a number of global climate zones have been classified ranging from large-scale primary climate zones (tropical, sub-tropical, temperate, sub-polar, polar) to much higher-resolution, regional climate zones (e.g., the Köppen-Geiger classification, Kottek et al. 2006). Biomes are adapted to regional climates (Figure 2.1) and may shift as climate, land surface characteristics (e.g., geomorphology, hydrology), CO₂ fertilisation, and fire interact. These biomes and processes therein are subject to modes of natural variability in the ocean-atmosphere system that result in regionally wetter/drier or hotter/cooler periods having temporal scales from weeks to months (e.g., Southern Annular Mode), months to seasons (e.g., Madden-Julian Oscillation), years (e.g., El Niño Southern Oscillation) and decades (e.g. Pacific Decadal Oscillation). Furthermore, climate and weather extremes (such as drought, heat waves, very heavy rainfall, strong winds), whose frequency, intensity and duration are often a function of large-scale modes of variability, impact ecosystems at various space and time scales.

It is *very likely* that changes to natural climate variability as a result of global warming has and will continue to impact terrestrial ecosystems with subsequent impacts on land processes (Hulme et al. 1999; Parmesan and Yohe 2003; Di Lorenzo et al. 2008; Kløve et al. 2014; Berg et al. 2015; Lemordant et al. 2016; Pecl et al. 2017). This chapter assesses climate variability and change, particularly extreme weather and climate, in the context of desertification, land degradation, food security and terrestrial ecosystems more generally. This section does specifically assess the impacts of climate variability and climate change on desertification, land degradation and food security as these impacts are assessed respectively in Chapters 3, 4 and 5. This chapter begins with an assessment of observed warming on land.

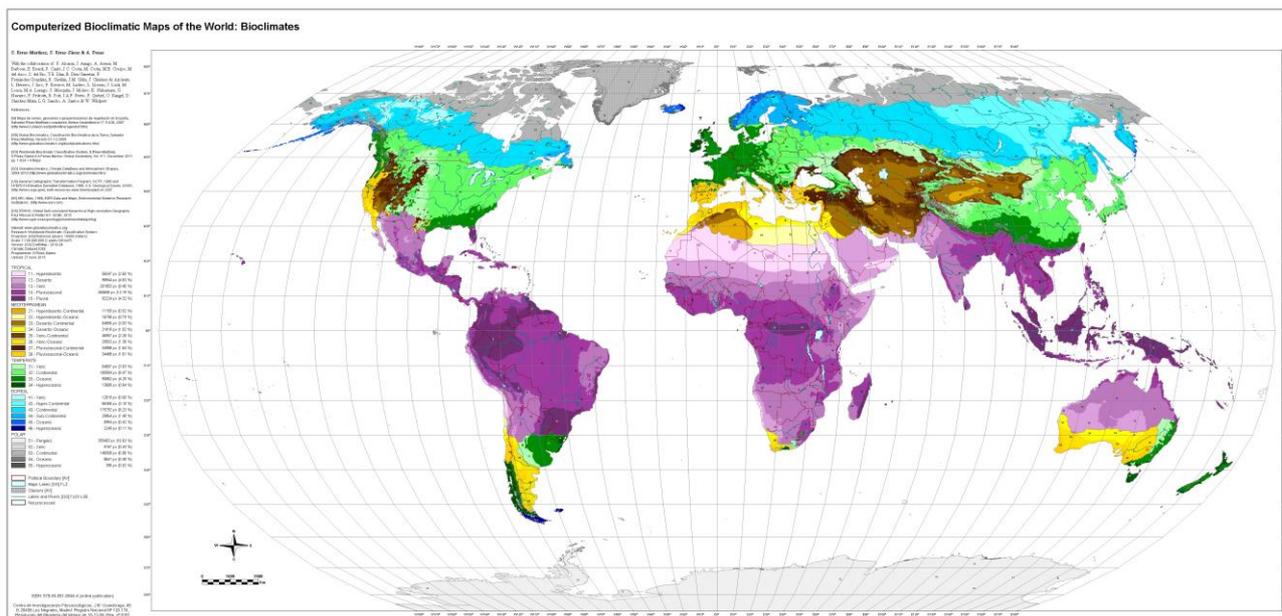


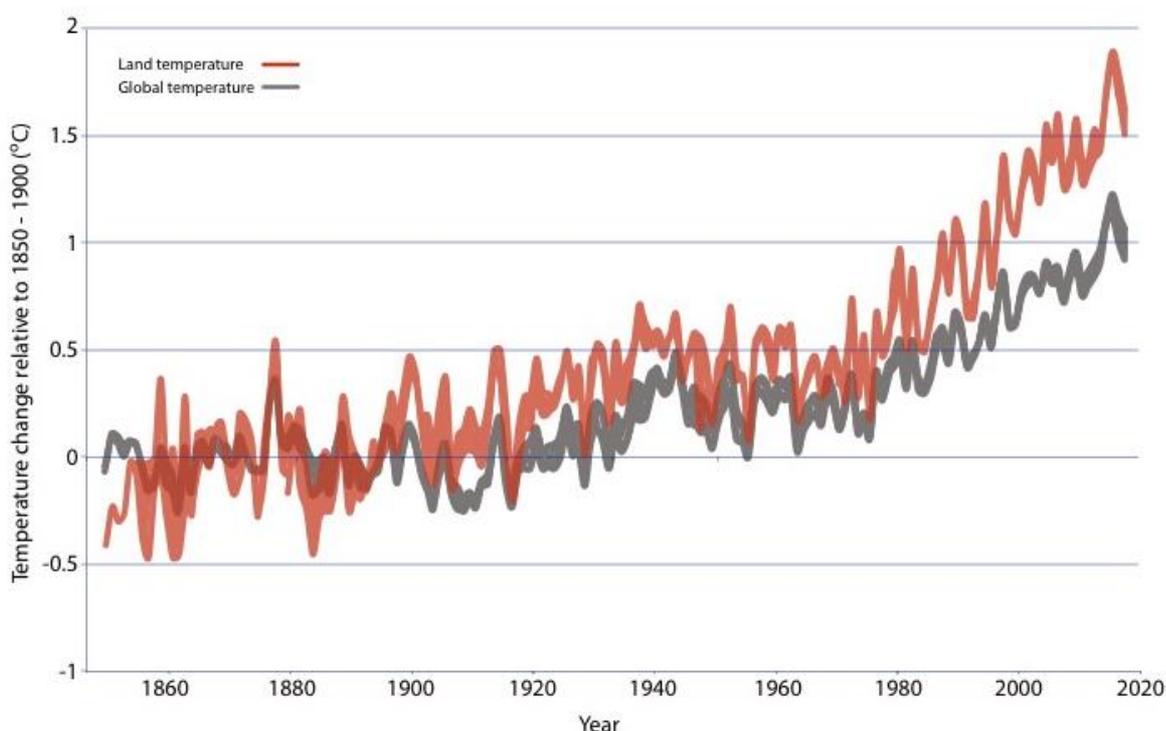
Figure 2.1 Worldwide Bioclimatic Classification System, 1996-2018. After (Rivas-Martinez et al. 2011).
Online at <http://www.globalbioclimatics.org>.

2.2.1.2 Changes in global land surface air temperature

Based on analysis of several global and regional land surface air temperature (LSAT) datasets, AR5 concluded that the global LSAT had increased over the instrumental period of record, with the warming rate approximately double that reported over the oceans since 1979 and that “it is certain that globally averaged LSAT has risen since the late 19th century and that this warming has been particularly marked since the

1 1970s". Warming found in the global land datasets is also in a broad agreement with station observations
 2 (Hartmann et al. 2013a).

3
 4 Since AR5, LSAT datasets have been improved and extended. The National Center for Environmental
 5 Information, which is a part of the US National Oceanic and Atmospheric Administration (NOAA),
 6 developed a new version of the Global Historical Climatology Network (GHCNm, version 4) dataset. The
 7 dataset provides an expanded set of station temperature records with more than 25,000 total monthly
 8 temperature stations compared to 7200 in versions v2 and v3 (Menne et al. 2018). Goddard Institute for
 9 Space Studies, which is a part of the US National Aeronautics and Space Administration, (NASA/GISS)
 10 provides estimate of land and ocean temperature anomalies (GISTEMP). The GISTEMP land temperature
 11 anomalies are based upon primarily NOAA/GHCN version 3 dataset (Lawrimore et al. 2011) and account for
 12 urban effects through nightlight adjustments (Hansen et al. 2010). The Climatic Research Unit of the
 13 University of East Anglia, UK (CRUTEM) dataset, now version CRUTEM4.6, incorporates additional
 14 stations (Jones et al. 2012). Finally, the Berkeley Earth Surface Temperature (BEST) dataset provides LSAT
 15 from 1750 to present based on almost 46,000 time series and has the longest temporal coverage of the four
 16 datasets (Rohde et al. 2013). This dataset was derived with methods distinct from those used for
 17 development of the NOAA and NASA datasets and the CRU dataset.
 18
 19



20
 21
 22 **Figure 2.2 Evolution of land surface air temperature (LSAT) and global mean surface temperature**
 23 **(GMST) over the period of instrumental observations. Red line shows annual mean LSAT in the**
 24 **Berkeley, CRUTEM4, GHCNv4 and GISTEMP datasets, expressed as departures from global average**
 25 **LSAT in 1850–1900, with the red line thickness indicating inter-dataset range. Gray shaded line shows**
 26 **annual mean Global Mean Surface Temperature (GMST) in the HadCRUT4, NOAA Global Temp,**
 27 **GISTEMP and Cowtan&Way datasets (monthly values of which were reported in the Special Report on**
 28 **Global Warming of 1.5°C (Allen et al. 2018)).**
 29

30 According to the available observations in the four datasets, the globally averaged LSAT increased by
 31 1.44°C from the preindustrial period (1850–1900) to present (1999–2018). The warming from the late 19th
 32 century (1881–1900) to present (1999–2018) was 1.41°C (1.31°C–1.51°C) (Table 2.1). The 1.31°C–1.51°C
 33 range represents the spread in median estimates from the four available land datasets and does not reflect
 34 uncertainty in data coverage or methods used. Based on the Berkeley dataset (the longest dataset with the
 35 most extensive land coverage) the total increase in LSAT between the average of the 1850–1900 period and

1 the 1999–2018 period was 1.52°C, (1.39°C–1.66°C; 95% confidence).

2 The extended and improved land datasets reaffirmed the AR5 conclusion that it is certain that globally
3 averaged LSAT has risen since the preindustrial period and that this warming has been particularly marked
4 since the 1970s (Figure 2.2).

5
6 **Table 2.1 Increases in land surface air temperature (LSAT) from preindustrial
7 period and the late 19th century to present (1999–2018).**

| Reference period | Dataset of LSAT increase (C°) | | | |
|--------------------------|-------------------------------|---------|-----------|---------|
| | Berkeley | CRUTEM4 | GHCNm, v4 | GISTEMP |
| <i>Preindustrial</i> | 1.52 | 1.31 | NA | NA |
| 1850–1900 | 1.39–1.66 (95% confidence) | | | |
| <i>Late 19th century</i> | 1.51 | 1.31 | 1.37 | 1.45 |
| 1881–1900 | 1.40–1.63 (95% confidence) | | | |

8
9 Recent analyses of LSAT and sea surface temperature (SST) observations as well as analyses of climate
10 model simulations have refined our understanding of underlying mechanisms responsible for a faster rate of
11 warming over land than over oceans. Analyses of paleo records, historical observations, model simulations,
12 and underlying physical principles are all in agreement that that land is warming faster than the oceans as a
13 result of differences in evaporation, land-climate feedbacks (Section 2.5), and changes in the aerosol forcing
14 over land (Braconnot et al. 2012; Joshi et al. 2013; Sejas et al. 2014; Byrne and O’Gorman 2013, 2015;
15 Wallace and Joshi 2018; Allen et al. 2019) (*very high confidence*). There is also *high confidence* that
16 difference in land and ocean heat capacity is not the primary reason for a faster land than ocean warming.
17 For the recent period the land-to-ocean warming ratio is in close agreement between different observational
18 records (about 1.6) and the CMIP5 climate model simulations (the *likely* range of 1.54°C to 1.81°C). Earlier
19 studies analysing slab ocean models (models in which it is assumed that the deep ocean has equilibrated)
20 produced a higher land temperature increases than sea surface temperature (Manabe et al. 1991; Sutton et al.
21 2007).

22
23 It is certain that globally averaged LSAT has risen faster than GMST from preindustrial (1850–1900) to
24 present day (1999–2018). This is because the warming rate of the land compared to the ocean is substantially
25 higher over the historical period (by approximately 60%) and because Earth surface is approximately one
26 third land and two thirds ocean. This enhanced land warming impacts land processes with implications for
27 desertification (Section 2.2.2, Chapter 3), food security (Section 2.2.3, Chapter 5), terrestrial ecosystems
28 (Section 2.2.4), and GHG and non-GHG fluxes between the land and climate (Section 2.3, Section 2.4).
29 Future changes in land characteristics through adaptation and mitigation processes and associated land-
30 climate feedbacks can dampen warming in some regions and enhance warming in others (Section 2.5).

31 32 **2.2.2 Climate driven changes in aridity**

33
34 Desertification is defined and discussed at length in Chapter 3 of this report and is a function of both human
35 activity and climate variability and change. There are uncertainties in distinguishing between historical
36 climate-caused aridification and desertification and also future projections of aridity as different
37 measurement methods of aridity do not agree on historical or projected changes (3.1.1, 3.2.1). However,
38 warming trends over drylands are twice the global average (Lickley and Solomon 2018) some temperate
39 drylands are projected to convert to subtropical drylands as a result of an increased drought frequency
40 causing reduced soil moisture availability in the growing season (Engelbrecht et al. 2015; Schlaepfer et al.
41 2017). We therefore assess with *medium confidence* that a warming climate will result in regional increases
42 in the spatial extent of drylands under mid- and high emission scenarios and that these regions will warm
43 faster than the global average warming rate.

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40
41
42
43
44
45
46
47
48
49

2.2.3 The influence of climate change on food security

Food security and the various components thereof is addressed in depth in Chapter 5. Climate variables relevant to food security and food systems are predominantly temperature and precipitation-related, but also include integrated metrics that combine these and other variables (like solar radiation, wind, humidity) and extreme weather and climate events including storm surge (see 5.2.1). The impact of climate change through changes in these variables is projected to negatively impact all aspects of food security (food availability, access, utilisation and stability), leading to complex impacts on global food security (Chapter 5) (Table 5.1) (*high confidence*).

Climate change will have regionally distributed impacts, even under aggressive mitigation scenarios (Howden et al. 2007; Rosenzweig et al. 2013; Challinor et al. 2014; Parry et al. 2005; Lobell and Tebaldi 2014; Wheeler and Von Braun 2013). For example, in the northern hemisphere the northward expansion of warmer temperatures in the middle and higher latitudes will lengthen the growing season (Gregory and Marshall 2012; Yang et al. 2015b) which may benefit crop productivity (Parry et al. 2004; Rosenzweig et al., 2014; Deryng et al. 2016). However, continued rising temperatures are expected to impact global wheat yields by about 4–6% reductions for every degree of temperature rise (Liu et al. 2016a; Asseng et al. 2015) and across both mid- and low latitude regions, rising temperatures are also expected to be a constraining factor for maize productivity by the end of the century (Bassu et al. 2014; Zhao et al. 2017). Although there has been a general reduction in frost occurrence during winter and spring and a lengthening of the frost free season in response to growing concentrations of greenhouse gases (Fischer and Knutti 2014; Wypych et al. 2017), there are regions where the frost season length has increased e.g. southern Australia (Crimp et al. 2016). Despite the general reduced frost season length, late spring frosts may increase risk of damage to warming induced precocious vegetation growth and flowering (Meier et al. 2018). Observed and projected warmer minimum temperatures have and will continue to reduce the number of winter chill units required by particularly fruit crops (Luedeling 2012). Crop yields are impacted negatively by increases of seasonal rainfall variability in the tropics, sub-tropics, water-limited and high elevation environments, drought severity and growing season temperatures have a negative impact on crop yield (IFPRI 2009; Schlenker and Lobell 2010; Müller et al. 2017; Parry et al. 2004; Wheeler and Von Braun 2013; Challinor et al. 2014).

Changes in extreme weather and climate (Section 2.2.5) have negative impacts on food security through regional reductions of crop yields. A recent study shows that between 18-43% of the explained yield variance of four crops (maize, soybeans, rice and spring wheat) is attributable to extremes of temperature and rainfall, depending on the crop type (Vogel et al. 2019). Climate shocks, particularly severe drought impact low-income small-holder producers disproportionately (Vermeulen et al. 2012b; Rivera Ferre 2014). Extremes also compromise critical food supply chain infrastructure, making the transport and access to harvested food more difficult (Brown et al. 2015; Fanzo et al. 2018). There is *high confidence* that the impacts of enhanced climate extremes, together with non-climate factors such as nutrient limitation, soil health and competitive plant species, generally outweighs the regionally positive impacts of warming (Lobell et al. 2011; Leakey et al. 2012; Porter et al. 2014; Gray et al. 2016; Pugh et al. 2016; Wheeler and Von Braun 2013; Beer 2018).

2.2.4 Climate-driven changes in terrestrial ecosystems

Previously, the IPCC AR5 reported high confidence that the Earth's biota composition and ecosystem processes have been strongly affected by past changes in global climate, but the rates of the historic climate change are lower than those projected for the 21st century under high warming scenarios like RCP8.5

1 (Settele et al. 2015a). There is *high confidence* that as a result of climate changes over recent decades many
2 plant and animal species have experienced range size and location changes, altered abundances, and shifts in
3 seasonal activities (Urban 2015a; Ernakovich et al. 2014; Elsen and Tingley 2015; Hatfield and Prueger
4 2015; Urban 2015b; Savage and Vellend 2015; Yin et al. 2016; Pecl et al. 2017; Gonsamo et al. 2017;
5 Fadrique et al. 2018; Laurance et al. 2018). There is high confidence that climate zones have already shifted
6 in many parts of the world primarily as an increase of dry, arid climates accompanied by a decrease of polar
7 climates (Chan and Wu 2015; Chen and Chen 2013; Spinoni et al. 2015b). Regional climate zones shifts
8 have been observed over the Asian monsoon region (Son and Bae 2015), Europe (Jylhä et al. 2010), China
9 (Yin et al. 2019), Pakistan (Adnan et al. 2017), the Alps (Rubel et al. 2017) and North-Eastern Brazil,
10 Southern Argentina, the Sahel, Zambia and Zimbabwe, the Mediterranean area, Alaska, Canada and North-
11 Eastern Russia (Spinoni et al. 2015b).

12
13 There is *high confidence* that bioclimates zones will further shift as the climate warms (Williams et al. 2007;
14 Rubel and Kottek 2010; Garcia et al. 2016; Mahony et al. 2017; Law et al. 2018). There is also high
15 confidence that novel, unprecedented climates (climate conditions with no analog in the observational
16 record) will emerge, particularly the tropics (Williams and Jackson 2007; Colwell et al. 2008a; Mora et al.
17 2013, 2014; Hawkins et al. 2014; Mahony et al. 2017; Maule et al. 2017). It is *very likely* that terrestrial
18 ecosystems and land processes will be exposed to disturbances beyond the range of current natural
19 variability as a result of global warming, even under low- to medium-range warming scenarios, and these
20 disturbances will alter the structure, composition and functioning of the system (Settele et al. 2015b;
21 Gauthier et al. 2015; Seddon et al. 2016).

22
23 In a warming climate many species will be unable to track their climate niche as it moves, especially those in
24 extensive flat landscapes with low dispersal capacity and in the tropics whose thermal optimum is already
25 near current temperature (Diffenbaugh and Field 2013; Warszawski et al. 2013). Range expansion in higher
26 latitudes and elevations as a result of warming often, but not exclusively occurs in abandoned lands (Harsch
27 et al. 2009; Landh usser et al. 2010; Gottfried et al. 2012; Boisvert-Marsh et al. 2014; Bryn and Potthoff
28 2018; Rumpf et al. 2018; Buitenwerf et al. 2018; Steinbauer et al. 2018). This expansion typically favours
29 thermophilic species at the expense of cold adapted species as climate becomes suitable for lower
30 latitude/altitude species (Rumpf et al. 2018). In temperate drylands, however, range expansion can be
31 countered by intense and frequent drought conditions which result in accelerated rates of taxonomic change
32 and spatial heterogeneity in an ecotone (Tietjen et al. 2017).

33
34 Since the advent of satellite observation platforms, a global increase in vegetation photosynthetic activity
35 (i.e. greening) as evidenced through remotely sensed indices such as leaf area index (LAI) and normalised
36 difference vegetation index (NDVI). Three satellite-based leaf area index (GIMMS3g, GLASS and
37 GLOMAP) records imply increased growing season LAI (greening) over 25–50% and browning over less
38 than 4% of the global vegetated area, resulting in greening trend of $0.068 \pm 0.045 \text{ m}^2 \text{ m}^{-2} \text{ yr}^{-1}$ over 1982–2009
39 (Cao et al. 2016). Greening has been observed in southern Amazonia, southern Australia, the Sahel and
40 central Africa, India, eastern China and the northern extratropical latitudes (Myneni et al. 1997; de Jong et al.
41 2012; Los 2013; Piao et al. 2015; Mao et al. 2016; Zhu et al. 2016; Carlson et al. 2017; Forzieri et al. 2017;
42 Pan et al. 2018; Chen et al. 2019). Greening has been attributed to direct factors, namely human land use
43 management and indirect factors such as CO₂ fertilisation, climate change, nitrogen deposition (Donohue et
44 al. 2013; Keenan et al. 2016; Zhu et al. 2016). Indirect factors have been used to explain most greening
45 trends primarily through CO₂ fertilisation in the tropics and through an extended growing season and
46 increased growing season temperatures as a result of climate change in the high latitudes (Fensholt et al.
47 2012; Zhu et al. 2016). The extension of the growing season in high latitudes has occurred together with an
48 earlier spring greenup (the time at which plants begin to produce leaves in northern mid- and high-latitude
49 ecosystems) (Goetz et al. 2015; Xu et al. 2016a, 2018) with subsequent earlier spring carbon uptake (2.3
50 days per decade) and gross primary productivity (GPP) (Pulliainen et al. 2017). The role of direct factors of

1 greening are being increasingly investigated and a recent study has attributed over a third of observed global
2 greening between 2000 and 2017 to direct factors, namely afforestation and croplands, in China and India
3 (Chen et al. 2019).

4
5 It should be noted that as measured greening is a product of satellite-derived radiance data, and as such does
6 not provide information on ecosystem health indicators such as species composition and richness,
7 homeostasis, absence of disease, vigor, system resilience and the different components of ecosystems
8 (Jørgensen et al. 2016). For example, a regional greening attributable to croplands expansion or
9 intensification might occur at the expense of ecosystem biodiversity.

10
11 Within the global greening trend are also detected regional decreases in vegetation photosynthetic activity
12 (i.e. browning) in northern Eurasia, the southwestern USA, boreal forests in North America, Inner Asia and
13 the Congo Basin, largely as a result of intensified drought stress. Since the late-1990s rates and extents of
14 browning have exceeded those of greening in some regions, the collective result of which has been a
15 slowdown of the global greening rate (de Jong et al. 2012; Pan et al. 2018). Within these long-term trends,
16 interannual variability of regional greening and browning is attributable to regional climate variability,
17 responses to extremes such as drought, disease and insect infestation and large-scale teleconnective controls
18 such as ENSO and the Atlantic Multi-decadal Organization (Verbyla 2008; Revadekar et al. 2012; Epstein et
19 al. 2018; Zhao et al. 2018).

20
21 Projected increases in drought conditions in many regions suggest long-term global vegetation greening
22 trends are at risk of reversal to browning in a warmer climate (de Jong et al. 2012; Pan et al. 2018; Pausas
23 and Millán 2018). On the other hand, in higher latitudes vegetation productivity is projected to increase as a
24 result of higher atmospheric CO₂ concentrations and longer growing periods as a result of warming here (Ito
25 et al. 2016)(Section 2.3, Box 2.3). Additionally, climate-driven transitions of ecosystems, particularly range
26 changes, can take years to decades for the equilibrium state to be realised and the rates of these “committed
27 ecosystem changes” (Jones et al. 2009) vary between low and high latitudes (Jones et al. 2010). Furthermore,
28 as direct factors are poorly integrated into Earth systems models (ESMs) uncertainties in projected trends of
29 greening and browning are further compounded (Buitenwerf et al. 2018; Chen et al. 2019). Therefore, there
30 is *low confidence* in the projection of global greening and browning trends.

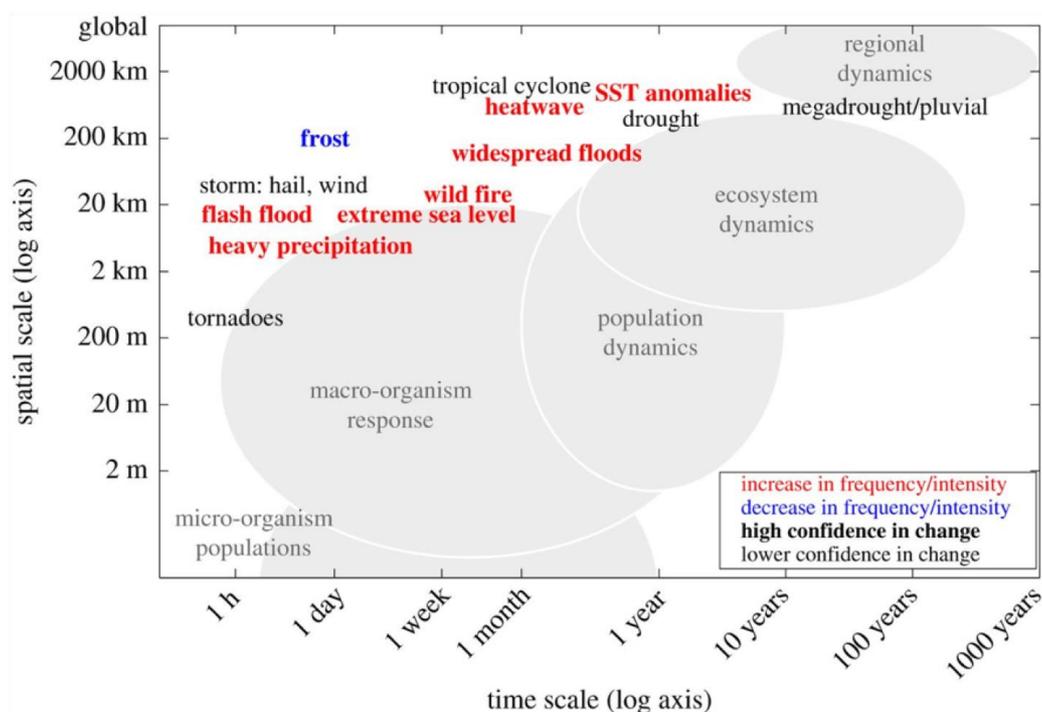
31
32 Increased atmospheric CO₂ concentrations have both direct and indirect effects on terrestrial ecosystems (see
33 Sections 2.2.2 and 2.2.3, Box 2.3). The direct effect is primarily through increased vegetation photosynthetic
34 activity as described above. Indirect effects include decreased evapotranspiration that may offset the
35 projected impact of drought in some water-stressed plants through improved water use efficiency in
36 temperate regions suggesting that some rain-fed cropping systems and grasslands will benefit from elevated
37 atmospheric CO₂ concentrations (Roy et al. 2016a; Milly and Dunne 2016; Swann et al. 2016; Chang et al.
38 2017; Zhu et al. 2017). In tropical regions increased flowering activity is associated primarily with
39 increasing atmospheric CO₂ suggesting a long-term increase in flowering activity may persist in some
40 vegetation, particularly mid-story trees and tropical shrubs, and enhance reproduction levels until limited by
41 nutrient availability or climate factors like drought frequency, rising temperatures and reduced insolation
42 (Pau et al. 2018).

43 44 **2.2.5 Climate extremes and their impact on land functioning**

45
46 Extreme weather events are generally defined as the upper or lower statistical tails of the observed range of
47 values of climate variables or climate indicators (e.g., temperature/rainfall or drought/aridity indices
48 respectively). Previous IPCC reports have reported with *high confidence* on the increase of many types of
49 observed extreme temperature events (Seneviratne et al. 2012b; Hartmann et al. 2013b; Hoegh-Guldberg et
50 al. 2018). However, as a result of observational constraints, increases in precipitation extremes are less

1 confident, except in observation rich regions with dense, long-lived station networks such as Europe and
 2 North America where there have been likely increases in the frequency or intensity of heavy rainfall.

3
 4 Extreme events occur across a wide range of time and space scales (Figure 2.3) and may include individual,
 5 relatively short-lived weather events (e.g., extreme thunderstorms storms) or a combination or accumulation
 6 of non-extreme events (Colwell et al. 2008b; Kundzewicz and Germany 2012) e.g., moderate rainfall in a
 7 saturated catchment having the flood peak at mean high tide (Leonard et al. 2014). Combinatory processes
 8 leading to a significant impact are referred to as a compound event and are a function of the nature and
 9 number of physical climate and land variables, biological agents such as pests and disease, the range of
 10 spatial and temporal scales, the strength of dependence between processes, and the perspective of the
 11 stakeholder who defines the impact (Leonard et al. 2014; Millar and Stephenson 2015). Current *confidence*
 12 in the impact of compound events on land is *low* as the multi-disciplinary approaches needed to address the
 13 problem are few (Zscheischler et al. 2018) and the rarity of compound extreme climatic events renders the
 14 analysis of impacts difficult.



16
 17
 18 **Figure 2.3** Spatial and temporal scales of typical extreme weather and climate events and the biological
 19 systems they impact (shaded grey). Individuals, populations and ecosystems within these space-time
 20 ranges respond to relevant climate stressors. Red (blue) labels indicate an increase (decrease) in the
 21 frequency or intensity of the event, with bold font reflecting confidence in the change. Non-bold black
 22 labels indicate low confidence in observed changes in frequency or intensity of these events. Each event
 23 type indicated in the figure is likely to affect biological systems at all temporal and spatial scales located
 24 to the left and below the specific event position in the figure. From Ummerhofer and Meehl (2017).

25 26 2.2.5.1 Changes in extreme temperatures, heat waves and drought

27 It is *very likely* that most land areas have experienced a decrease in the number of cold days and nights, and
 28 an increase in the number of warm days and unusually hot nights (Orlowsky and Seneviratne 2012;
 29 Seneviratne et al. 2012; Mishra et al. 2015; Ye et al. 2018). Although there is no consensus definition of heat
 30 waves as some heat wave indices have relative thresholds and others absolute thresholds, trends between
 31 indices of the same type show that recent heat-related events have been made more frequent or more intense
 32 due to anthropogenic greenhouse gas emissions in most land regions (Lewis and Karoly 2013; Smith et al.
 33 2013b; Scherer and Diffenbaugh 2014; Fischer and Knutti 2015; Ceccherini et al. 2016; King et al. 2016;

1 Bador et al. 2016; Stott et al. 2016; King 2017; Hoegh-Guldberg et al. 2018). Globally, 50–80 % of the land
2 fraction is projected to experience significantly more intense hot extremes than historically recorded (Fischer
3 and Knutti 2014; Diffenbaugh et al. 2015; Seneviratne et al. 2016). There is *high confidence* that heat waves
4 will increase in frequency, intensity and duration into the 21st century (Russo et al. 2016; Ceccherini et al.
5 2017; Herrera-Estrada and Sheffield 2017) and under high emission scenarios heat waves by the end of the
6 century may become extremely long (more than 60 consecutive days) and frequent (once every two years) in
7 Europe, North America, South America, Africa, Indonesia, the Middle East, south and south east Asia and
8 Australia (Rusticucci 2012; Cowan et al. 2014; Russo et al. 2014; Scherer and Diffenbaugh 2014; Pal and
9 Eltahir 2016; Rusticucci et al. 2016; Schär 2016; Teng et al. 2016; Dosio 2017; Mora et al. 2017; Dosio et al.
10 2018; Lehner et al. 2018; Lhotka et al. 2018; Lopez et al. 2018; Tabari and Willems 2018). Furthermore,
11 unusual heat wave conditions today will occur regularly by 2040 under the RCP 8.5 scenario (Russo et al.
12 2016). The intensity of heat events may be modulated by the land cover and soil characteristics (Miralles et
13 al. 2014; Lemordant et al. 2016; Ramarao et al. 2016). Where temperature increase results in decreased soil
14 moisture, latent heat flux is reduced while sensible heat fluxes is increased allowing surface air temperature
15 to rise further. However, this feedback may be diminished if the land surface is irrigated through the
16 enhanced evapotranspiration (Mueller et al. 2015; Siebert et al. 2017)(Section 2.5.2.2).

17
18 Drought (Qin et al. 2013), including megadroughts of the last century, e.g., the Dustbowl drought (Hegerl et
19 al. 2018), Chapter 5), is a normal component of climate variability (Hoerling et al. 2010; Dai 2011) and may
20 be seasonal, multi-year (Van Dijk et al. 2013) or multi-decadal (Hulme 2001) with increasing degrees of
21 impact on the regional activities. This interannual variability is controlled particularly through remote sea
22 surface temperature (SST) forcings such as the Inter-decadal Pacific Oscillation (IPO) and the Atlantic
23 Multi-decadal Oscillation (AMO), El Niño/Southern Oscillation (ENSO) and Indian Ocean Dipole (IOD),
24 that cause drought as a result of reduced rainfall (Kelley et al. 2015; Dai 2011; Hoell et al. 2017; Espinoza et
25 al. 2018). In some cases however, large scale SST modes do not fully explain the severity of drought some
26 recent event attribution studies have identified a climate change fingerprint in several regional droughts, e.g.,
27 the western Amazon (Erfanian et al. 2017), southern Africa (Funk et al. 2018; Yuan et al. 2018), southern
28 Europe and the Mediterranean including North Africa (Kelley et al. 2015; Wilcox et al. 2018), parts of North
29 America (Williams et al. 2015; Mote et al. 2016), Russia (Otto et al. 2012), India (Ramarao et al. 2015) and
30 Australia (Lewis and Karoly 2013).

31
32 Long-term global trends in drought are difficult to determine because of this natural variability, potential
33 deficiencies in drought indices (especially in how evapotranspiration is treated) and the quality and
34 availability of precipitation data (Sheffield et al. 2012; Dai 2013; Trenberth et al. 2014; Nicholls and
35 Seneviratne 2015; Mukherjee et al. 2018). However, regional trends in frequency and intensity of drought
36 are evident in several parts of the world, particularly in low latitude land areas, such as the Mediterranean,
37 North Africa and Middle East (Vicente-Serrano et al. 2014; Spinoni et al. 2015a; Dai and Zhao 2017; Páscoa
38 et al. 2017), many regions of sub-Saharan Africa (Masih et al. 2014; Dai and Zhao 2017), Central China
39 (Wang et al. 2017e), the southern Amazon (Fu et al. 2013; Espinoza et al. 2018), India (Ramarao et al.
40 2016), east and south Asia, parts of North America and eastern Australia (Dai and Zhao 2017). A recent
41 analysis of 4500 meteorological droughts globally found increased drought frequency over the U.S. East
42 Coast, Amazonia and north-eastern Brazil, Patagonia, the Mediterranean region, most of Africa and north-
43 eastern China with decreased drought frequency over northern Argentina, Uruguay and northern Europe
44 (Spinoni et al. 2019). The study also found drought intensity has become more severe over north-western
45 U.S., parts of Patagonia and southern Chile, the Sahel, the Congo River basin, southern Europe, north-
46 eastern China, and south-eastern Australia, whereas the eastern U.S., south-eastern Brazil, northern Europe,
47 and central-northern Australia experienced less severe droughts. In addition to the IPCC SR15 assessment of
48 medium confidence in increased drying over the Mediterranean region (Hoegh-Guldberg et al. 2018), It is
49 further assessed with *medium confidence* an increased frequency and intensity of drought in Amazonia and
50 north-eastern Brazil, Patagonia, most of Africa and north-eastern China.

1 There is *low confidence* on how large-scale modes of variability will respond to a warming climate (Deser et al. 2012; Liu 2012; Christensen et al. 2013; Hegerl et al. 2015; Newman et al. 2016). Although, there is evidence for an increased frequency of extreme ENSO events, such as the 1997/98 El Niño and 1988/89 La Niña (Cai et al. 2014a, 2015) and extreme positive phases of the IOD (Christensen et al. 2013; Cai et al. 2014b). However, The assessment by the SR15 was retained on an increased regional drought risk (*medium confidence*), specifically over the Mediterranean and South Africa at both 1.5°C and 2°C warming levels compared to present day, with drought risk at 2°C being significantly higher than at 1.5°C (Hoegh-Guldberg et al. 2018).

10 **2.2.5.2 Impacts of heat extremes and drought on land**

11 There is *high confidence* that heat extremes such as unusually hot nights, extremely high daytime temperatures, heat waves and drought are damaging to crop production (Chapter 5). Extreme heat events impact a wide variety of tree functions including reduced photosynthesis, increased photooxidative stress, leaves abscise, a decreased growth rate of remaining leaves and decreased growth of the whole tree (Teskey et al. 2015). Although trees are more resilient to heat stress than grasslands (Teuling et al. 2010), it has been observed that different types of forest (e.g., needleleaf vs. broadleaf) respond differently to drought and heat waves (Babst et al. 2012). For example, in the Turkish Anatolian forests net primary productivity (NPP) generally decreased during drought and heat waves events between 2000 and 2010 but in a few other regions, NPP of needle leaf forests increased (Erşahin et al. 2016). However, forests may become less resilient to heat stress in future due to the long recovery period required to replace lost biomass and the projected increased frequency of heat and drought events (Frank et al. 2015a; McDowell and Allen 2015; Johnstone et al. 2016; Stevens-Rumann et al. 2018). Additionally, widespread regional tree mortality may be triggered directly by drought and heat stress (including warm winters) and exacerbated by insect outbreak and fire (Neuvonen et al. 1999; Breshears et al. 2005; Berg et al. 2006; Soja et al. 2007; Kurz et al. 2008b; Allen et al. 2010) .

27 Gross primary production (GPP) and soil respiration form the first and second largest carbon fluxes from terrestrial ecosystems to the atmosphere in the global carbon cycle (Beer et al. 2010; Bond-Lamberty and Thomson 2010). Heat extremes impact the carbon cycle through altering these and change ecosystem-atmosphere CO₂ fluxes and the ecosystem carbon balance. Compound heat and drought events result in a stronger carbon sink reduction compared to single-factor extremes as GPP is strongly reduced and ecosystem respiration less so (Reichstein et al. 2013; Von Buttlar et al. 2018). In forest biomes, however, GPP may increase temporarily as a result of increased insolation and photosynthetic activity as was seen during the 2015-2016 ENSO related drought over Amazonia (Zhu et al. 2018). Longer extreme events (heat wave or drought or both) result in a greater reduction in carbon sequestration and may also reverse long-term carbon sinks (Ciais et al. 2005; Phillips et al. 2009; Wolf et al. 2016b; Ummenhofer and Meehl 2017; Von Buttlar et al. 2018; Reichstein et al. 2013). Furthermore, extreme heat events may impact the carbon cycle beyond the lifetime of the event. These lagged effects can slow down or accelerate the carbon cycle: it will slow down if reduced vegetation productivity and/or widespread mortality after an extreme drought are not compensated by regeneration, or speed up if productive tree and shrub seedlings cause rapid regrowth after windthrow or fire (Frank et al. 2015a). Although some ecosystems may demonstrate resilience to a single heat climate stressor like drought, e.g. forests, compound effects of, e.g., deforestation, fire and drought potentially can result in changes to regional precipitation patterns and river discharge, losses of carbon storage and a transition to a disturbance-dominated regime (Davidson et al. 2012). Additionally, adaptation to seasonal drought may be overwhelmed by multi-year drought and their legacy effects (Brando et al. 2008; da Costa et al. 2010).

48 Under medium and high emission scenarios, global warming will exacerbate heat stress thereby amplifying deficits in soil moisture and runoff despite uncertain precipitation changes (Ficklin and Novick 2017; Berg and Sheffield 2018; Cook et al. 2018; Dai et al. 2018; Engelbrecht et al. 2015; Ramarao et al. 2015; Grillakis

2019). This will increase the rate of drying causing drought to set in quicker, become more intense and widespread, last longer and could result in an increased global aridity (Dai 2011; Prudhomme et al. 2014).

The projected changes in the frequency and intensity of extreme temperatures and drought is expected to result in decreased carbon sequestration by ecosystems and degradation of ecosystems health and loss of resilience (Trumbore et al. 2015). Also affected are many aspects of land functioning and type including agricultural productivity (Lesk et al. 2016a), hydrology (Mosley 2015; Van Loon and Laaha 2015), vegetation productivity and distribution (Xu et al. 2011; Zhou et al. 2014), carbon fluxes and stocks and other biogeochemical cycles (Frank et al. 2015b; Doughty et al. 2015; Schlesinger et al. 2016). Carbon stocks are particularly vulnerable to extreme events due to their large carbon pools and fluxes, potentially large lagged impacts and long recovery times to regain lost stocks (Frank et al. 2015a)(Section 2.2).

2.2.5.3 *Changes in heavy precipitation*

A large number of extreme rainfall events have been documented over the past decades (Coumou and Rahmstorf 2012; Seneviratne et al. 2012a; Trenberth 2012; Westra et al. 2013; Espinoza et al. 2014; Guhathakurta et al. 2017; Taylor et al. 2017; Thompson et al. 2017; Zilli et al. 2017). The observed shift in the trend distribution of precipitation extremes is more distinct than for annual mean precipitation and the global land fraction experiencing more intense precipitation events is larger than expected from internal variability (Fischer and Knutti 2014; Espinoza et al. 2018; Fischer et al. 2013) . As a result of global warming the number of record-breaking rainfall events globally has increased significantly by 12% during the period 1981 to 2010 compared to those expected due to natural multi-decadal climate variability (Lehmann et al. 2015) and the IPCC SR15 reports robust increases in observed precipitation extremes for annual maximum 1-day precipitation (RX1day) and consecutive 5-day precipitation (RX5day) (Hoegh-Guldberg et al. 2018; Schleussner et al. 2017). A number of extreme rainfall events have been attributed to human influence (Min et al. 2011; Pall et al. 2011; Sippel and Otto 2014; Trenberth et al. 2015; Krishnan et al. 2016) and the largest fraction of anthropogenic influence is evident in the most rare and extreme events (Fischer and Knutti 2014).

A warming climate is expected to intensify the hydrological cycle as a warmer climate facilitates more water vapour in the atmosphere, as approximated by the Clausius-Clapeyron (C-C) relationship, with subsequent effects on regional extreme precipitation events (Christensen and Christensen 2003; Pall et al. 2007; Berg et al. 2013; Wu et al. 2013; Guhathakurta et al. 2017; Thompson et al. 2017; Taylor et al. 2017; Zilli et al. 2017)(Manola et al. 2018). Furthermore, changes to the dynamics of the atmosphere amplify or weaken future precipitation extremes at the regional scale (O’Gorman 2015; Pfahl et al. 2017). Continued anthropogenic warming is very likely to increase the frequency and intensity of extreme rainfall in many regions of the globe (Seneviratne et al. 2012a; Mohan and Rajeevan 2017; Prein et al. 2017; Stott et al. 2016) although many GCMs underestimate observed increased trends in heavy precipitation suggesting a substantially stronger intensification of future heavy rainfall than the multi-model mean (Borodina et al. 2017; Min et al. 2011). Furthermore, the response of extreme convective precipitation to warming remains uncertain because GCMs and regional climate models (RCMs) are unable to explicitly simulate sub-grid scale processes such as convection, the hydrological cycle and surface fluxes and have to rely on parameterisation schemes for this (Crétat et al. 2012; Rossow et al. 2013; Wehner 2013; Kooperman et al. 2014; O’Gorman 2015; Larsen et al. 2016; Chawla et al. 2018; Kooperman et al. 2018; Maher et al. 2018; Rowell and Chadwick 2018). High-resolution regional climate models that explicitly resolve convection have a better representation of extreme precipitation but are dependent on the GCM to capture large scale environment in which the extreme event may occur (Ban et al. 2015; Prein et al. 2015; Kendon et al. 2017) . Interannual variability of precipitation extremes in the convective tropics are not well captured by global models (Allan and Liu 2018).

There is low confidence in the detection of long-term observed and projected seasonal and daily trends of extreme snowfall. The narrow rain–snow transition temperature range at which extreme snowfall can occur

1 that is relatively insensitive to climate warming and subsequent large interdecadal variability (Kunkel et al.
2 2013; O’Gorman 2014, 2015).

3 4 **2.2.5.4 Impacts of precipitation extremes on different land cover types**

5 More intense rainfall leads to water redistribution between surface and ground water in catchments as water
6 storage in the soil decreases (green water) and runoff and reservoir inflow increases (blue water) (Liu and
7 Yang 2010; Eekhout et al. 2018). This results in increased surface flooding and soil erosion, increased plant
8 water stress and reduced water security, which in terms of agriculture means an increased dependency on
9 irrigation and reservoir storage (Nainggolan et al. 2012; Favis-Mortlock and Mullen 2011; García-Ruiz et al.
10 2011; Li and Fang 2016; Chagas and Chaffe 2018). As there is high confidence of a positive correlation
11 between global warming and future flood risk, land cover and processes are likely to be negatively impacted,
12 particularly near rivers and in floodplains (Kundzewicz et al. 2014; Alfieri et al. 2016; Winsemius et al.
13 2016; Arnell and Gosling 2016; Alfieri et al. 2017; Wobus et al. 2017).

14
15 In agricultural systems heavy precipitation and inundation can delay planting, increases soil compaction, and
16 causes crop losses through anoxia and root diseases (Posthumus et al. 2009). In tropical regions flooding
17 associated with tropical cyclones can lead to crop failure from both rainfall and storm surge. In some cases
18 flooding can affect yield more than drought, particularly in tropical regions (e.g. India) and in some mid/high
19 latitude regions such as China and central and northern Europe (Zampieri et al. 2017). Waterlogging of
20 croplands and soil erosion also negatively affect farm operations and block important transport routes (Vogel
21 and Meyer 2018; Kundzewicz and Germany 2012). Flooding can be beneficial in drylands if the floodwaters
22 infiltrate and recharge alluvial aquifers along ephemeral river pathways, extending water availability into dry
23 seasons and drought years and support riparian systems and human communities (Kundzewicz and Germany
24 2012; Guan et al. 2015). Globally, the impact of rainfall extremes on agriculture is less than that of
25 temperature extremes and drought, although in some regions and for some crops, extreme precipitation
26 explains a greater component of yield variability, e.g. of maize in the Mid-Western USA and southern Africa
27 (Ray et al. 2015; Lesk et al. 2016b; Vogel et al. 2019) .

28
29 Although many soils on floodplains regularly suffer from inundation, the increases in the magnitude of flood
30 events means that new areas with no recent history of flooding are now becoming severely affected (Yellen
31 et al. 2014). Surface flooding and associated soil saturation often results in decreased soil quality through
32 nutrient loss, reduced plant productivity, stimulates microbial growth and microbial community composition,
33 negatively impacts soil redox and increases greenhouse gas emissions (Bossio and Scow 1998; Niu et al.
34 2014; Barnes et al. 2018; Sánchez-Rodríguez et al. 2019). The impact of flooding on soil quality is
35 influenced by management systems that may mitigate or exacerbate the impact. Although soils tend to
36 recover quickly after floodwater removal, the impact of repeated extreme flood events over longer timescales
37 on soil quality and function is unclear (Sánchez-Rodríguez et al. 2017).

38
39 Flooding in ecosystems may be detrimental through erosion or permanent habitat loss, or beneficial, as a
40 flood pulse brings nutrients to downstream regions (Kundzewicz et al. 2014). Riparian forests can be
41 damaged through flooding; however, increased flooding may also be of benefit to forests where upstream
42 water demand has lowered stream flow, but this is difficult to assess and the effect of flooding on forests is
43 not well studied (Kramer et al. 2008; Pawson et al. 2013). Forests may mitigate flooding, however flood
44 mitigation potential is limited by soil saturation and rainfall intensity (Pilaš et al. 2011; Ellison et al. 2017b).
45 Some grassland species under heavy rainfall and soil saturated conditions responded negatively with
46 decreased reproductive biomass and germination rates (Gellesch et al. 2017), however overall productivity in
47 grasslands remains constant in response to heavy rainfall (Grant et al. 2014).

48
49 Extreme rainfall alters responses of soil CO₂ fluxes and CO₂ uptake by plants within ecosystems and
50 therefore result in changes in ecosystem carbon cycling (Fay et al. 2008; Frank et al. 2015a). Extreme
51 rainfall and flooding limits oxygen in soil which may suppress the activities of soil microbes and plant roots

1 and lower soil respiration and therefore carbon cycling (Knapp et al. 2008; Rich and Watt 2013; Philben et
2 al. 2015). However, the impact of extreme rainfall on carbon fluxes in different biomes differs. For example,
3 extreme rainfall in mesic biomes reduces soil CO₂ flux to the atmosphere and GPP whereas in xeric biomes
4 the opposite is true, largely as a result of increased soil water availability (Knapp and Smith 2001; Heisler
5 and Knapp 2008; Heisler-White et al. 2009; Zeppel et al. 2014; Xu and Wang 2016; Liu et al. 2017b; Connor
6 and Hawkes 2018).

7
8 As shown above greenhouse gas fluxes between the land and atmosphere are affected by climate. The next
9 section assesses these fluxes in greater detail and the potential for land as a carbon sink.

13 **Cross-Chapter Box 3: Fire and Climate Change**

14
15 Raman Sukumar (India), Almut Arneth (Germany), Werner Kurz (Canada), Andrey Sirin (Russian
16 Federation), Louis Verchot (Colombia/The United States of America)

17
18 Fires have been a natural part of Earth's geological past and its biological evolution since at least the late
19 Silurian, about 400 million years ago (Scott 2000). Presently, roughly 3% of the Earth's land surface burns
20 annually which affects both energy and matter exchanges between the land and atmosphere (Stanne et al.
21 2009). Climate is a major determinant of fire regimes through its control of fire weather, as well as through
22 its interaction with vegetation productivity (fuel availability) and structure (fuel distribution and
23 flammability) (Archibald et al. 2013) at the global (Krawchuk and Moritz 2011), regional (Pausas and Paula
24 2012) and local landscape (Mondal and Sukumar 2016) scales. Presently, humans are the main cause of fire
25 ignition with lightning playing a lesser role globally (Bowman et al. 2017; Harris et al. 2016), although the
26 latter factor has been predominantly responsible for large fires in regions such as the North American boreal
27 forests (Veraverbeke et al. 2017). Humans also influence fires by actively extinguishing them, reducing
28 spread and managing fuels.

29 *Historical trends and drivers in land area burnt*

30
31 While precipitation has been the major influence on wildfire regimes in pre-Industrial times, human
32 activities have become the dominant drivers since then. There was less biomass burning during the 20th
33 century than at any time during the past two millennia as inferred from charcoal sedimentary records (Doerr
34 and Santín 2016), though there has been an increase in the most recent decades (Marlon et al. 2016). Trends
35 in land area burnt have varied regionally (Giglio et al. 2013). Northern Hemisphere Africa has experienced a
36 fire decrease of 1.7 Mha yr⁻¹ (-1.4% yr⁻¹) since 2000, while Southern Hemisphere Africa saw an increase of
37 2.3 Mha yr⁻¹ (+1.8% yr⁻¹) during the same period. Southeast Asia witnessed a small increase of 0.2 Mha yr⁻¹
38 (+2.5% yr⁻¹) since 1997, while Australia experienced a sharp decrease of about 5.5 Mha yr⁻¹ (-10.7% yr⁻¹)
39 during 2001–2011, followed by an upsurge in 2011 that exceeded the annual area burned in the previous 14
40 years. A recent analysis using the Global Fire Emissions Database v.4 (GFED4s) that includes small fires
41 concluded that the net reduction in land area burnt globally during 1998–2015 was -24.3±8.8% (-
42 1.35±0.49% yr⁻¹) (Andela et al. 2017). However, from the point of fire emissions it is important to consider
43 the land cover types which have experienced changes in area burned; in this instance, most of the declines
44 have come from grasslands, savannas and other non-forest land cover types (Andela et al. 2017). Significant
45 increases in forest area burned (with higher fuel consumption per unit area) have been recorded in western
46 and boreal North America (Abatzoglou and Williams 2016; Ansmann et al. 2018) and in boreal Siberia
47 (Ponomarev et al. 2016) in recent times. The 2017 and 2018 fires in British Columbia, Canada, were the
48 largest ever recorded since the 1950s with 1.2 Mha and 1.4 Mha of forest burnt, respectively (Hanes et al.
49 2018) and smoke from these fires reaching the stratosphere over central Europe (Ansmann et al. 2018).

50
51 Climate variability and extreme climatic events such as severe drought, especially those associated with the

1 El Niño Southern Oscillation (ENSO), play a major role in fire upsurges as in equatorial Asia (Huijnen et al.
2 2016). Fire emissions in tropical forests increased by 133% on average during and following six El Niño
3 years compared to six La Niña years during 1997–2016, due to reductions in precipitation and terrestrial
4 water storage (Chen et al. 2017). The expansion of agriculture and deforestation in the humid tropics has also
5 made these regions more vulnerable to drought-driven fires (Davidson et al. 2012; Brando et al. 2014). Even
6 when deforestation rates were overall declining, as in the Brazilian Amazon during 2003–2015, the incidence
7 of fire increased by 36% during the drought of 2015 (Aragão et al. 2018).

9 ***GHG emissions from fires***

10 Emissions from wildfires and biomass burning are a significant source of greenhouse gases (CO₂, CH₄,
11 N₂O), carbon monoxide (CO), carbonaceous aerosols, and an array of other gases including non-methane
12 volatile organic compounds (NMVOC) (Akagi et al. 2011; Van Der Werf et al. 2010). GFED4s has updated
13 fire-related carbon emission estimates biome-wise, regionally and globally, using higher resolution input
14 data gridded at 0.25°, a new burned area dataset with small fires, improved fire emission factors (Akagi et al.
15 2011; Urbanski 2014) and better fire severity characterisation of boreal forests (van der Werf et al. 2017).
16 The estimates for the period 1997–2016 are 2.2 GtC yr⁻¹, being highest in the 1997 El Niño (3.0 GtC yr⁻¹)
17 and lowest in 2013 (1.8 GtC yr⁻¹). Furthermore, fire emissions during 1997–2016 were dominated by
18 savanna (65.3%), followed by tropical forest (15.1%), boreal forest (7.4%), temperate forest (2.3%), peatland
19 (3.7%) and agricultural waste burning (6.3%) (van der Werf et al. 2017).

20
21 Fires not only transfer carbon from land to the atmosphere but also between different terrestrial pools: from
22 live to dead biomass to soil, including partially charred biomass, charcoal and soot constituting 0.12–0.39
23 GtC yr⁻¹ or 0.2–0.6% of annual terrestrial NPP (Doerr and Santín 2016). Carbon from the atmosphere is
24 sequestered back into regrowing vegetation at rates specific to the type of vegetation and other
25 environmental variables (Loehman et al. 2014). Fire emissions are thus not necessarily a net source of carbon
26 into the atmosphere, as post-fire recovery of vegetation can sequester a roughly equivalent amount back into
27 biomass over a time period of one to a few years (in grasslands and agricultural lands) to decades (in forests)
28 (Landry and Matthews 2016). Fires from deforestation (for land use change) and on peatlands (which store
29 more carbon than terrestrial vegetation) obviously are a net source of carbon from the land to the atmosphere
30 (Turetsky et al. 2014); these types of fires were estimated to emit 0.4 GtC yr⁻¹ in recent decades (van der
31 Werf et al. 2017). Peatland fires dominated by smouldering combustion under low temperatures and high
32 moisture conditions can burn for long periods (Turetsky et al. 2014).

34 ***Fires, land degradation/desertification and land-atmosphere exchanges***

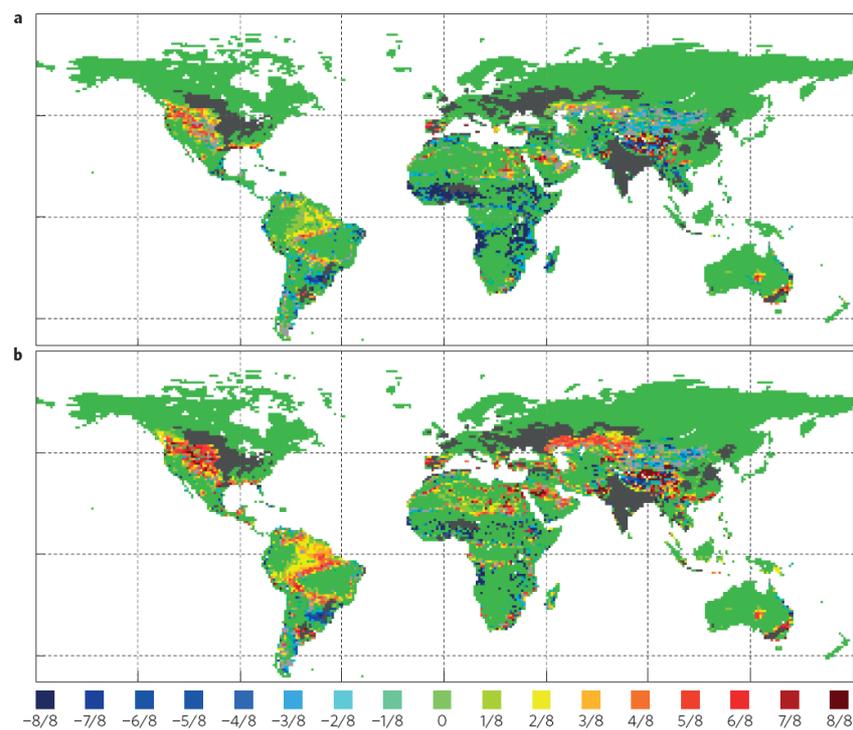
35 Flammable ecosystems are generally adapted to their specific fire regimes (Bond et al. 2005). A fire regime
36 shift alters vegetation and soil properties in complex ways, both in the short- and the long-term, with
37 consequences for carbon stock changes, albedo, fire-atmosphere-vegetation feedbacks and the ultimate
38 biological capacity of the burnt land (Bond et al. 2004; Bremer and Ham 1999; MacDermott et al. 2016;
39 Tepley et al. 2018; Moody et al. 2013; Veraverbeke et al. 2012) A fire-driven shift in vegetation from a
40 forested state to an alternative stable state such as a grassland (Fletcher et al. 2014; Moritz 2015) with much
41 less carbon stock is a distinct possibility. Fires cause soil erosion through action of wind and water (Moody
42 et al. 2013) thus resulting in land degradation (see Chapter 4) and eventually desertification (see Chapter 3).
43 Fires also affect carbon exchange between land and atmosphere through ozone (retards photosynthesis) and
44 aerosol (slightly increases diffuse radiation) emissions; the net effect on global GPP during 2002–2011 is
45 estimated to be -0.86 ± 0.74 GtC yr⁻¹ (Yue and Unger 2018).

47 ***Fires under future climate change***

48 Temperature increase and precipitation decline would be the major driver of fire regimes under future
49 climates as evapotranspiration increases and soil moisture decreases (Pechony and Shindell 2010; Aldersley
50 et al. 2011; Abatzoglou and Williams 2016; Fernandes et al. 2017). The risk of wildfires in future could be
51 expected to change, increasing significantly in North America, South America, central Asia, southern

1 Europe, southern Africa, and Australia (Liu et al. 2010). There is emerging evidence that recent regional
2 surges in wildland fires are being driven by changing weather extremes, thereby signalling geographical
3 shifts in fire proneness (Jolly et al. 2015). Fire weather season has already lengthened by 18.7% globally
4 between 1979 and 2013, with statistically significant increases across 25.3% but decreases only across
5 10.7% of Earth's land surface covered with vegetation; even sharper changes have been observed during the
6 second half of this period (Jolly et al. 2015). Correspondingly, the global area experiencing long fire weather
7 season (defined as experiencing fire weather season greater than one standard deviation (SD) from the mean
8 global value) has increased by 3.1% per annum or 108.1% during 1979–2013. Fire frequencies under 2050
9 conditions are projected to increase by approximately 27% globally, relative to the 2000 levels, with changes
10 in future fire meteorology playing the most important role in enhancing global wildfires, followed by land
11 cover changes, lightning activities and land use, while changes in population density exhibit the opposite
12 effects (Huang et al. 2014).

13
14 However, climate is only one driver of a complex set of environmental, ecological and human factors in
15 influencing fire regimes (Bowman et al. 2011a). While these factors lead to complex projections of future
16 burnt area and fire emissions (Knorr et al. 2016a,b), human exposure to wildland fires could still increase
17 because of population expansion into areas already under high risk of fires (Knorr et al. 2016a,b). There are
18 still major challenges in projecting future fire regimes, and how climate, vegetation and socio/economic
19 factors will interact (Hantson et al. 2016; Harris et al. 2016). There is also need for integrating various fire
20 management strategies, such as fuel-reduction treatments in natural and planted forests, with other
21 environmental and societal considerations to achieve the goals of carbon emissions reductions, maintain
22 water quality, biodiversity conservation and human safety (Moritz et al. 2014; Gharun et al. 2017).



23
24

Cross-Chapter Box 3, Figure 1: The probability of low-fire regions becoming fire prone (positive values), or of fire-prone areas changing to a low-fire state (negative values) between 1971–2000 and 2017–2100 based on eight-Earth system model (ESM) ensembles, two Shared Socio-economic Pathways (SSPs; see (Jiang 2014)) and two Representative Concentration Pathways (RCPs). Light grey: areas where at least one ensemble simulation predicts a positive and one a negative change (lack of agreement). Dark grey: area with >50% past or future cropland. Fire-prone areas are defined as having a fire frequency of >0.01 yr⁻¹ a RCP4.5 emissions with SSP3 demographics. b, RCP8.5 emissions with SSP5 demographics (Knorr et al. 2016a)

In summary, climate change is playing an increasing role in determining wildfire regimes along-side human activity (*medium confidence*), with future climate variability expected to enhance the risk and severity of wildfires in many biomes such as tropical rainforests (*high confidence*). Fire weather seasons have lengthened globally between 1979 and 2013 (*low confidence*). Global land area burned has declined in recent decades, mainly due to less burning in grasslands and savannas (*high confidence*). While drought remains the dominant driver of fire emissions, there has recently been increased fire activity in some tropical and temperate regions during normal to wetter than average years due to warmer temperatures that increase vegetation flammability (*medium confidence*). The boreal zone is also experiencing larger and more frequent fires, and this may increase under a warmer climate (*medium confidence*)

2.3 Greenhouse gas fluxes between land and atmosphere

Land is simultaneously a source and sink for several greenhouse gases (GHGs). Moreover, both natural and anthropogenic processes determine fluxes of GHGs, making it difficult to separate “anthropogenic” and “non-anthropogenic” emissions and removals. A meeting report by the (IPCC 2010a) divided the processes responsible for fluxes from land into three categories: (1) the *direct effects* of anthropogenic activity due to changing land cover and land management; (2) the *indirect effects* of anthropogenic environmental change, such as climate change, carbon dioxide (CO₂) fertilisation, nitrogen deposition; and (3) *natural* climate variability and natural disturbances (e.g. wildfires, windrow, disease). The meeting report (IPCC 2010a) noted that it was impossible with any direct observation to separate direct anthropogenic effects from non-anthropogenic (indirect and natural) effects in the land sector. As a result, different approaches and methods for estimating the anthropogenic fluxes have been developed by different communities to suit their individual purposes, tools and data availability.

The major GHGs exchanged between land and the atmosphere discussed in this chapter are CO₂ (2.3.1), methane (CH₄, Section 2.3.2) and nitrous oxide (N₂O, Section 2.3.3). We estimate the total emissions from Agriculture, Forestry, and Other Land Use (AFOLU) to be responsible for approximately 22% of global anthropogenic GHG emissions over the period 2003-2012 (Smith et al. 2013a; Ciais et al. 2013a) (Table 2.2). The estimate is similar to that reported in AR5 (*high confidence*), with slightly more than half these emissions coming as non-CO₂ GHGs from agriculture. Emissions from AFOLU have remained relatively constant since AR4, although their relative contribution to anthropogenic emissions has decreased due to increases in emissions from the energy sector.

Table 2.2 Summary of average annual land use fluxes aggregated over the decades 2001 to 2010 and 2007 to 2016. We present averages to smooth the effects of inter-annual variability.

| Land use emissions | Mt | Gt CO ₂ e ¹ | Mt | Gt CO ₂ e |
|----------------------------------|-------------------------------------|-----------------------------------|-------------------------------------|----------------------|
| | CH ₄ or N ₂ O | | CH ₄ or N ₂ O | |
| | 2001-2010 | | 2007-2016 | |
| Land use CO ₂ | | | | |
| Bookkeeping model average | | 4.69 | | 4.84 |
| DGVM average | | 4.75 | | 4.70 |
| FAOSTAT | | 3.48 | | 2.81 |
| Non-CO ₂ GHGs | | | | |

| | | | | |
|--|--------|--------------|--------|--------------|
| Agricultural CH ₄ | | | | |
| FAOSTAT | 130.44 | 3.65 | 135.42 | 3.79 |
| USEPA ² | 141.64 | 3.97 | 147.83 | 4.14 |
| EDGAR ³ | 145.43 | 4.07 | 152.68 | 4.28 |
| Average | | 3.90 | | 3.97 |
| Agricultural N ₂ O | | | | |
| FAOSTAT | 6.83 | 1.81 | 7.34 | 1.95 |
| USEPA | 8.69 | 2.30 | 9.32 | 2.47 |
| EDGAR | 5.92 | 1.57 | 6.31 | 1.67 |
| Average | | 1.89 | | 2.21 |
| Total emissions from land use⁴ | | 10.48 | | 11.02 |
| Total emissions from all sources ⁵ | | 45.29 | | 50.72 |
| Land use emissions: Total emissions | | 23% | | 22% |

1 Data sources: Bookkeeping models: Hansis et al. 2015; Houghton & Nassikas 2017; DGVM average: Le Quéré et al.
2 2018; FAOSTAT: Tubiello et al. 2013; USEPA: USEPA 2012; EDGAR: Janssens-Maenhout et al. 2017.

3 ¹ All values expressed in units of CO₂eq are based on AR5 100 year Global Warming Potential values without climate-
4 carbon feedbacks (N₂O = 265; CH₄ = 28).

5 ² USEPA data are calculated from country data using IPCC Tier 1 methods (IPCC 2003) at 5-year increments through
6 2005 and are projected estimates from 2010 through 2030

7 ³ EDGAR data are complete only through 2012; 2007-2012 averages were computed in the second data column, for
8 comparison purposes only. They were not used in the calculation of the average fluxes for the 2007-2016 period.

9 ⁴ Total land use emissions were calculated as the sum of the average emissions from the bookkeeping models and the
10 average non-CO₂ GHG data from the different data sources (values in bold).

11 ⁵ Total emissions from all sources were calculated as the sum of total CO₂e emissions values for energy, industrial
12 sources, waste and other emissions from the PRIMAP database (Gütschow et al. 2016). They do not include emissions
13 from international bunkers and shipping.

16 2.3.1 Carbon Dioxide

17
18 This section is divided into four sub-sections (**Figure 2.4**): (1) the total net flux of CO₂ between land and
19 atmosphere, (2) the contributions of AFOLU¹ fluxes and the non-AFOLU land sink to that total net CO₂ flux,
20 (3) the gross emissions and removals comprising the net AFOLU flux, and (4) the gross emissions and
21 removals comprising the land sink. Emissions to the atmosphere are positive; removals from the atmosphere
22 are negative.

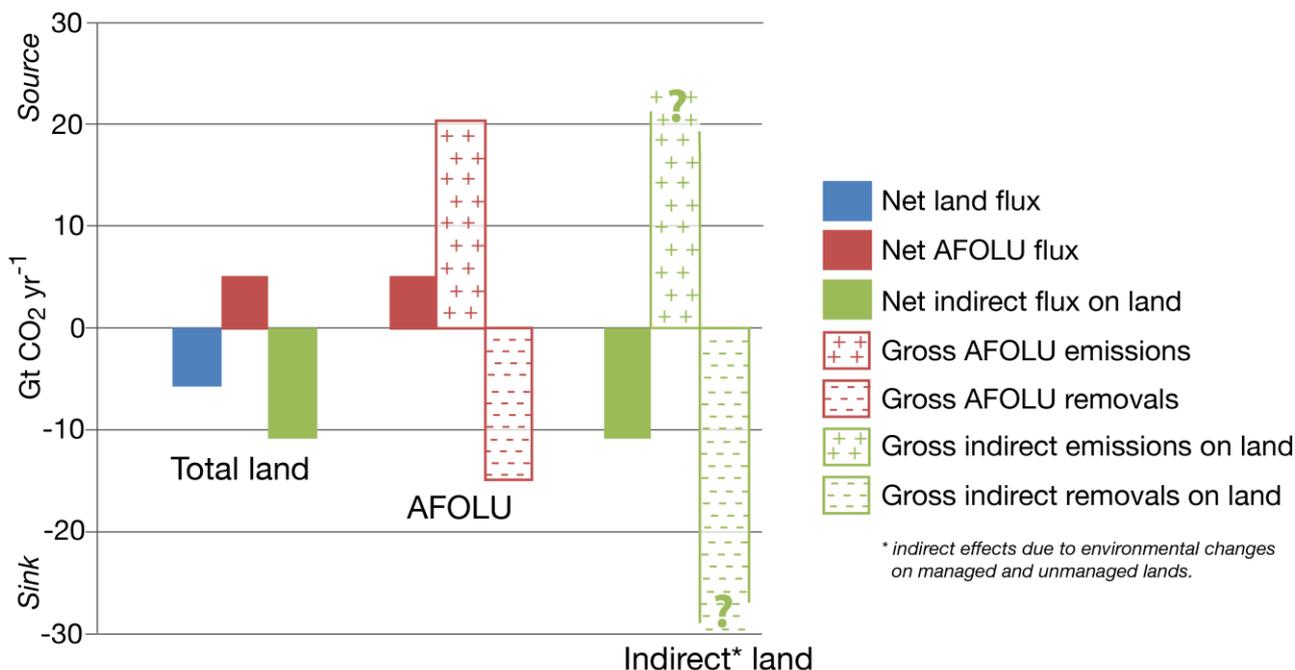
24 2.3.1.1 The total net flux of CO₂ between land and atmosphere

25 The net effects of all anthropogenic and non-anthropogenic processes on managed and unmanaged land
26 result in a net removal of CO₂ from the atmosphere (*high confidence*). This total net land-atmosphere
27 removal (defined here as *the total net land flux*) is estimated to have averaged 6.2 ± 2.6 GtCO₂ yr⁻¹ from
28 2007 to 2016 (**Table 2.3**.) The estimate is determined from summing the AFOLU and non-AFOLU fluxes
29 fluxes due to transient climate change, CO₂ fertilisation, nitrogen deposition calculated by models in the
30 global carbon budget (Le Quéré et al. 2018) and is consistent with inverse modelling techniques based on
31 atmospheric CO₂ concentrations and air transport (range: 5.1–8.8 GtCO₂ yr⁻¹) (Peylin et al. 2013; Van Der
32 Laan-Luijkx et al. 2017; Saeki and Patra 2017; Le Quéré et al. 2018) (See Box 2.2: for methods). A recent
33 inverse analysis, considering carbon transport in rivers and oceans, found a net flux of CO₂ for land within
34 this range, but a lower source from southern lands and a lower sink in northern lands (Resplandy et al. 2018).

35
36 The net removal of CO₂ by land has generally increased over the last 60 years in proportion to total

No FOOTNOTE/FOOTNOTE deleted

1 emissions of CO₂ (*high confidence*). Although land has been a net sink for CO₂ since around the middle of
 2 last century, it was a net source to the atmosphere before that time, primarily as a result of emissions from
 3 AFOLU (Le Quéré et al. 2018).
 4



5
 6
 7 **Figure 2.4 Net and gross fluxes of CO₂ from land (annual averages for 2008-2017).** [Left]: The total net flux
 8 of CO₂ between land and atmosphere (blue) is shown with its two component fluxes: Net AFOLU
 9 emissions (red) and the net land sink (green) due to indirect environmental effects and natural effects on
 10 managed and unmanaged lands. [Middle]: The gross emissions and removals contributing to the net
 11 AFOLU flux. [Right]: The gross emissions and removals contributing to the land sink.

12
 13 **2.3.1.2 Separation of the total net land flux into AFOLU fluxes and the land sink**

14 The total net flux of carbon between land and the atmosphere can be divided into fluxes due to direct human
 15 activities (i.e., AFOLU) and fluxes due to indirect anthropogenic and natural effects (i.e., the land sink)
 16 (Table 2.3). These two components are less certain than their sums, the total net flux of CO₂ between
 17 atmosphere and land. The land sink, estimated with DGVMs, is least certain (Figure 2.5).
 18

19 **Table 2.3 Perturbation of the global carbon cycle caused by anthropogenic activities (GtCO₂ yr⁻¹) (from (Le**
 20 **Quéré et al. 2018)).**

| | CO ₂ flux (GtCO ₂ yr ⁻¹), 10-year mean | | | | | |
|----------------------------------|--|------------|-------------|-------------|-------------|-------------|
| | 1960–1969 | 1970–1979 | 1980–1989 | 1990–1999 | 2000–2009 | 2008–2017 |
| Emissions | | | | | | |
| Fossil CO ₂ emissions | 11.4 ± 0.7 | 17.2 ± 0.7 | 19.8 ± 1.1 | 23.1 ± 1.1 | 28.6 ± 1.5 | 34. ± 1.8 |
| AFOLU net emissions | 5.5 ± 2.6 | 4.4 ± 2.6 | 4.4 ± 2.6 | 5.1 ± 2.6 | 4.8 ± 2.6 | 5.5 ± 2.6 |
| Partitioning | | | | | | |
| Growth in atmosphere | 6.2 ± 0.3 | 10.3 ± 0.3 | 12.5 ± 0.07 | 11.4 ± 0.07 | 14.7 ± 0.07 | 17.2 ± 0.07 |
| Ocean sink | 3.7 ± 1.8 | 4.8 ± 1.8 | 6.2 ± 1.8 | 7.3 ± 1.8 | 7.7 ± 1.8 | 8.8 ± 1.8 |
| Land sink | 4.4 ± 1.8 | 7.7 ± 1.5 | 6.6 ± 2.2 | 8.8 ± 1.8 | 9.9 ± 2.6 | 11.7 ± 2.6 |

(non-AFOLU)

| | | | | | | |
|---|------------|------------|------------|------------|------------|------------|
| Budget imbalance | 2.2 | -1.1 | -1.1 | 0.7 | 0.7 | 1.8 |
| Total net land flux (AFOLU - Land sink) | +1.1 ± 3.2 | -3.3 ± 3.0 | -2.2 ± 3.4 | -3.7 ± 2.2 | -5.1 ± 3.2 | -6.2 ± 3.7 |

2.3.1.2.1 Fluxes attributed to AFOLU

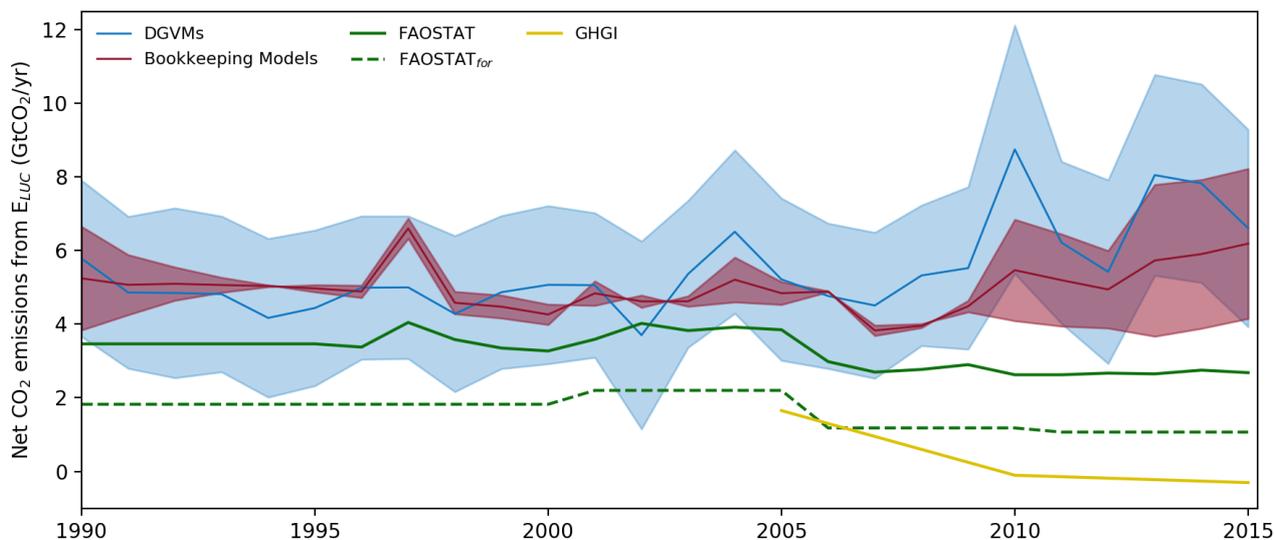
The modelled AFOLU flux was a net emission of 5.5 ± 2.6 GtCO₂ yr⁻¹ for 2008–2017, approximately 14% of total anthropogenic CO₂ emissions (Le Quéré et al. 2018, **Table 2.3**). This net flux was due to direct anthropogenic activities, predominately tropical deforestation, but also afforestation/reforestation, and fluxes due to forest management (e.g. wood harvest) and other types of land management, including agriculture, grasslands and scrub. The AFOLU flux is the mean of two estimates from bookkeeping models (Hansis et al. 2015; Houghton and Nassikas 2017), and this estimated mean is consistent with the mean obtained from an assemblage of Dynamic Global Vegetation Models (DGVMs) (Le Quéré et al. 2018, Box 2.2., **Figure 2.5**), although not all individual DGMVs include the same types of land use. Net CO₂ emissions from AFOLU have been relatively constant since 1900. AFOLU emissions were the dominant anthropogenic emissions until around the middle of the last century when fossil fuel emissions became dominant (Le Quéré et al. 2018). AFOLU activities have resulted in emissions of CO₂ over recent decades (*robust evidence, high agreement*) although there is a wide range of estimates from different methods and approaches (Smith et al. 2014; Houghton et al. 2012; Gasser and Ciais 2013; Pongratz et al. 2014; Tubiello et al. 2015; Grassi et al. 2018) (see Methods Box 2.2., **Figure 2.5** and **Figure 2.7**).

DGVMs and one bookkeeping model (Hansis et al. 2015) used spatially explicit, harmonised land-use change data (LUH2) (Hurtt et al. 2017) based on HYDE 3.2. The HYDE data, in turn, are based on changes in the areas of croplands and pastures. In contrast, the Houghton bookkeeping approach (Houghton and Nassikas 2017) used primarily changes in forest area from the FAO Forest Resource Assessment (FAO 2015a) and FAOSTAT to determine changes in land use. To the extent that forests are cleared for land uses other than crops and pastures, estimates from Houghton and Nassikas (2017, 2018) are higher than estimates from DGMVs. In addition, both bookkeeping models (Hansis et al. 2015; Houghton and Nassikas 2017) included estimates of carbon emissions in SE Asia from peat burning from the Global Fire Emissions Database (GFED version 4, (Randerson et al. 2015)) and from peat drainage (Hooijer et al. 2010).

Satellite-based estimates of CO₂ emissions from loss of tropical forests during 2000-2010 corroborate the modelled emissions but are quite variable: 4.8 GtCO₂ yr⁻¹ (Tyukavina et al. 2015), 3.0 GtCO₂ yr⁻¹ (Harris et al. 2015), 3.2 GtCO₂ yr⁻¹ (Achard et al. 2014) and 1.6 GtCO₂ yr⁻¹ (Baccini et al. 2017). Differences in estimates can be explained to a large extent by the different approaches used. For example, the analysis by (Tyukavina et al. 2015) led to a higher estimate because they used a finer spatial resolution. Three of the estimates considered losses in forest area and ignored degradation and regrowth of forests. Baccini et al. (2017) in contrast, included both losses and gains in forest area and losses and gains of carbon within forests (i.e., forest degradation and growth). The four remote sensing studies cited above also reported committed emissions; i.e., all of the carbon lost from deforestation was assumed to be released to the atmosphere in the year of deforestation. In reality, only some of the carbon in trees is not released immediately to the atmosphere at the time of deforestation. The unburned portion is transferred to woody debris and wood products. Both bookkeeping models and DGVMs account for the delayed emissions in growth and decomposition. Finally, the satellite-based estimates do not include changes in soil carbon.

In addition to differences in land-cover data sets between models and satellites, there are many other methodological reasons for differences (See Box 2.2:) (Houghton et al. 2012; Gasser and Ciais 2013; Pongratz et al. 2014; Tubiello et al. 2015). There are different definitions of land-cover type, including forest (e.g. FAO uses a tree cover threshold for forests of 10%; Tyukavina et al. (2017) used 25%), different estimates of biomass and soil carbon density (Mg C ha⁻¹), different approaches to tracking emissions through time (legacy effects), and different types of activity included (e.g. forest harvest, peatland drainage and fires). Most DGVMs only recently (since AR5) included forest management processes, such as tree harvesting and land clearing for shifting cultivation, leading to larger estimates of CO₂ emissions than when

1 these processes are not considered (Arneth et al. 2017; Erb et al. 2018). Grazing management has likewise
 2 been found to have large effects (Sanderman et al. 2017), and is not included in most DGVMs (Pugh et al.
 3 2015; Pongratz et al., 2018).



4
 5 **Figure 2.5** Global net CO₂ emissions due to AFOLU from different approaches (in GtCO₂ yr⁻¹). Red line:
 6 the mean and individual estimates from two bookkeeping models (Houghton and Nassikas 2017; Hansis
 7 et al. 2015). Blue line: the mean from DGVMs run with the same driving data with the pale blue shading
 8 showing the ±1 standard deviation range. Green line: data downloaded from FAOSTAT website
 9 (Tubiello et al. 2013); the dashed line is primarily forest-related emissions, while the solid green line also
 10 includes emissions from peat fires and peat draining. Yellow line: Greenhouse Gas Inventories (GHGI)
 11 based on country reports to UNFCCC (Grassi et al. 2018), data are shown only from 2005 because
 12 reporting in many developing countries became more consistent/reliable after this date. For more details
 13 on methods see Box 2.2:
 14

15 2.3.1.2.2 Nationally reported Greenhouse Gas Inventories (GHGI) values versus global model estimates

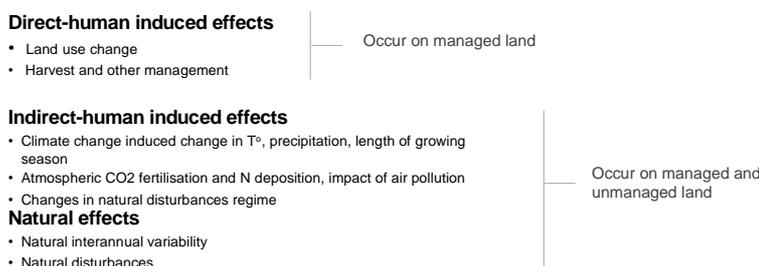
16 There are large differences globally (Figure 2.5), between estimates of net anthropogenic land-atmosphere
 17 fluxes of CO₂ from national GHGIs and from global models, and the same is true in many regions (Figure
 18 2.5). Fluxes reported to the UNFCCC through country GHGIs were noted as about 4.3 GtCO₂ yr⁻¹ lower
 19 (Grassi et al. 2018) than estimates from the bookkeeping model (Houghton et al. 2012a) used in the carbon
 20 budget for AR5 (Conway 2012; Ciais et al. 2013a). The anthropogenic emissions of CO₂ from AFOLU
 21 reported in countries' GHG inventories were 0.1 ± 1.0 GtCO₂ yr⁻¹ globally during 2005 to 2014 (*low*
 22 *confidence*) (Grassi et al. 2018) much lower than emission estimates from the two global bookkeeping
 23 models of 5.1 ± 2.6 GtCO₂ yr⁻¹ over the same time period (Quéré et al. 2018). Transparency and
 24 comparability in estimates can support measuring, reporting and verifying GHG fluxes under the UNFCCC,
 25 and also the global stocktake, which will assess globally the progress towards achieving the long-term goals
 26 of the Paris Agreement. These differences can be reconciled largely by taking account of the different
 27 approaches to defining anthropogenic in terms of different areas of land and treatment of indirect
 28 environmental change (Grassi et al. 2018).
 29

30 To date there has been one study that quantitatively reconciles the global model estimates with GHGIs
 31 (Grassi et al. 2018). The separation of anthropogenic from non-anthropogenic effects is impossible with
 32 direct observation (IPCC 2010a). The different approaches of models and GHGIs to estimating
 33 anthropogenic emissions and removals are shown in (Figure 2.6). The difficulty is that *indirect* effects of
 34 environmental changes (e.g. climate change and rising atmospheric CO₂) affect both managed and unmanaged
 35 lands, and some approaches treat these as anthropogenic while others do not. Bookkeeping models (e.g.
 36 Houghton and Nassikas 2017) attempt to estimate the fluxes of CO₂ driven by direct anthropogenic effects
 37 alone. DGVMs model the *indirect* environmental effects of climate and CO₂. If the indirect effects happen
 38 on land experiencing anthropogenic land cover change or management (harvest and regrowth), DGVMs treat
 39 this as anthropogenic. Country GHGIs separately report fluxes due to land conversion (e.g. forests to
 40 croplands) and fluxes due to land management (e.g. forest land remaining forest land). The “managed land
 41 proxy” is used as a pragmatic approach to estimate anthropogenic fluxes on managed lands, whereby

1 countries define the areas they consider managed, and include all of the emission and removals that occur on
 2 those lands. Emissions and removals are caused simultaneously by direct, indirect and natural drivers and are
 3 captured in the reporting, which often relies on inventories.
 4

5 Grassi et al. (2018) demonstrated that estimates of CO₂ emissions from global models and from nationally
 6 reported GHGIs were similar for deforestation and afforestation, but different for managed forests. Countries
 7 generally reported larger areas of managed forests than the models, and the carbon removals by these
 8 managed forests were also larger. The flux due to indirect effects on managed lands was quantified using
 9 post-processing of results from DGVMs, looking at the *indirect* effects of CO₂ and climate change on
 10 secondary forest areas. The derived DGVM *indirect* managed forest flux was found to account for most of
 11 the difference between the bookkeeping models and the inventories.
 12

a) Effects of various factors on the forest CO₂ fluxes and where they occur



b) Conceptual differences in defining the anthropogenic land CO₂ flux

IPCC AR5 and Global Carbon Budget:

Bookkeeping models
 "Land Use Change"

| | Managed land | Unmanaged land |
|--------------------------------|--------------|----------------|
| Direct human induced effects | ✓ | |
| Indirect human induced effects | | |
| Natural effects | | |

DGVMs:
 "Land Use Change" and "Land Sink"

| | Managed land | Unmanaged land |
|--------------------------------|--------------|----------------|
| Direct human induced effects | ✓ | |
| Indirect human induced effects | ✓ | ✓ |
| Natural effects | | |

Country GHG inventories:

"AFOLU (LULUCF)"

| | Managed land | Unmanaged land |
|--------------------------------|--------------|----------------|
| Direct human induced effects | ✓ | |
| Indirect human induced effects | ✓ | |
| Natural effects | ✓ | |

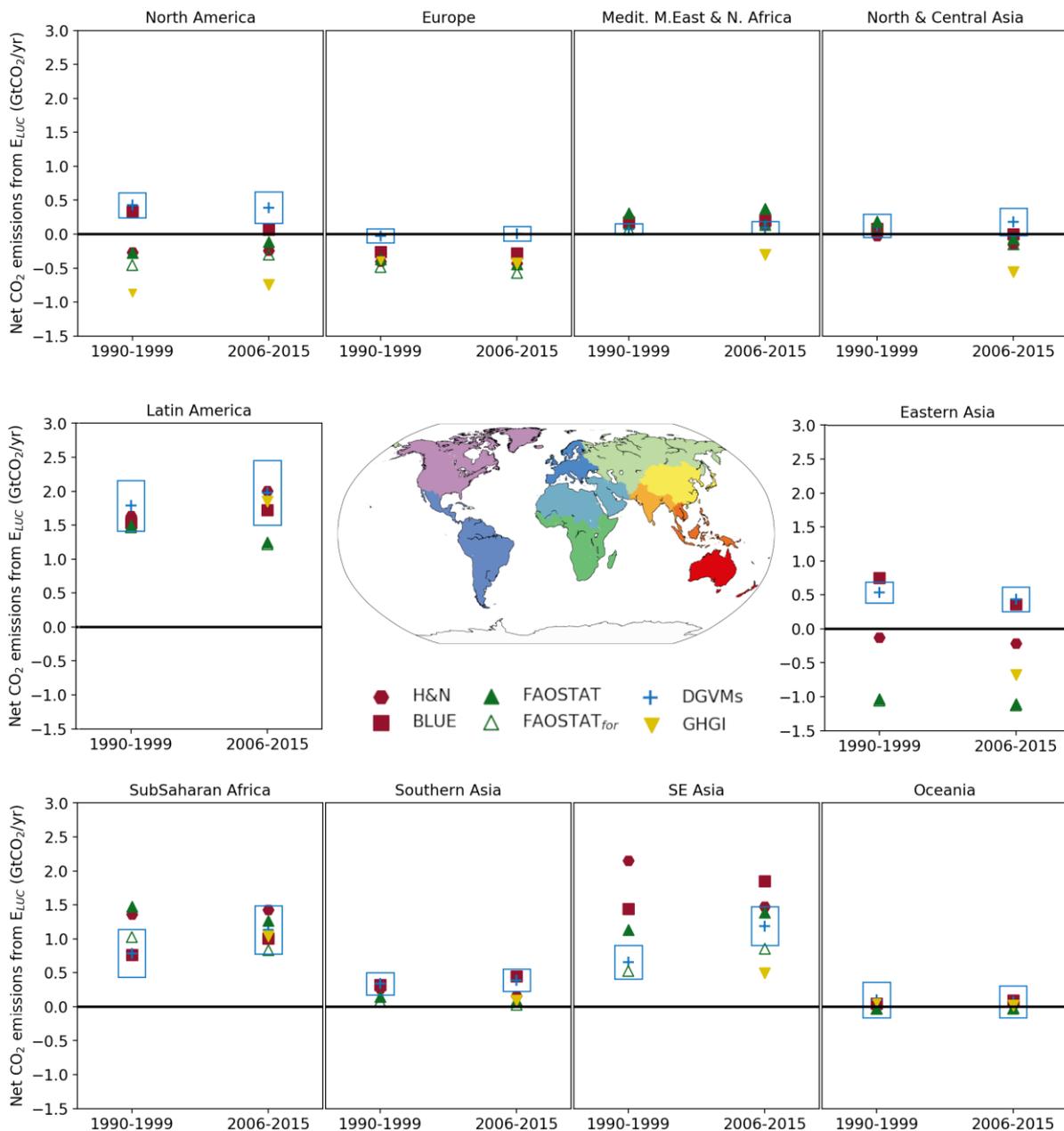
13
 14
 15 **Figure 2.6 Summary of the main conceptual differences between GHG Inventories and global models in**
 16 **considering what is the "anthropogenic land CO₂ flux". Adapted from Grassi et al. (2018): a) Effects of**
 17 **key processes on the land flux as defined by IPCC (2010) including where these effects occur (in**
 18 **managed and/or unmanaged lands); How these effects are captured in '(a) bookkeeping models that do**
 19 **not explicitly model the effects of environmental change, although some is implicitly captured in data on**
 20 **carbon densities and growth and decay rates; (b) Dynamic Global Vegetation Models (DGVMs) that**
 21 **include the effects of environmental change on all lands, and run the models with and without land use**
 22 **change to diagnose "land use change", the "land sink" is then conceptually assumed to be a natural**
 23 **response of land to the anthropogenic perturbation of environmental change, models include the effects**
 24 **of inter-annual climate variability, and some include fires but no other natural disturbances (c) GHG**
 25 **Inventories reported by countries to the UNFCCC that report all fluxes in areas the countries define as**
 26 **"managed land" but do not report unmanaged land. This is the CO₂ flux due to Land Use Land Use**
 27 **Change and Forestry (LULUCF) which is a part of the overall AFOLU (Agriculture, Forestry and**
 28 **Other Land Use) flux. The area of land considered as managed in the inventories is greater than that**
 29 **considered as subject to direct management activities (harvest and regrowth) in the models.**
 30

31 **2.3.1.2.3 Regional differences**

32 Figure 2.7 shows regional differences in emissions due to AFOLU. Recent increases in deforestation rates in
 33 some tropical countries have been partially balanced by increases in forest area in India, China, the USA and

1 Europe (FAO-FRA 2015). The trend in emissions from AFOLU since the 1990s is *uncertain* because some
 2 data suggest a declining rate of deforestation (FAO-FRA 2015), while data from satellites suggest an
 3 increasing rate (Kim 2014; Hansen et al. 2012). The disagreement results in part from differences in the
 4 definition of forest and approaches to estimating deforestation. The FAO defines deforestation as the
 5 conversion of forest to another land use (FAO-FRA 2015), while the measurement of forest loss by satellite
 6 may include wood harvests (forests remaining forests) and natural disturbances that are not directly caused
 7 by anthropogenic activity (e.g., forest mortality from droughts and fires). Trends in anthropogenic and
 8 natural disturbances may be in opposite directions. For example, recent drought-induced fires in the Amazon
 9 have increased the emissions from wildfires at the same time that emissions from anthropogenic
 10 deforestation have declined (Aragão et al. 2018). Furthermore, there have been advances since AR5 in
 11 estimating the GHG effects of different types of forest management (e.g. (Valade et al. 2017). Overall, there
 12 is *robust evidence and high agreement* for a net loss of forest area and tree cover in the tropics and a net
 13 gain, mainly of secondary forests and sustainably managed forests, in the temperate and boreal zones
 14 (Chapter 1).

15



16
 17
 18

Figure 2.7 Regional trends in net anthropogenic land-atmosphere CO₂ flux from a range of different approaches (in GtCO₂ yr⁻¹). Red symbols: bookkeeping models (hexagon - Houghton and Nassikas 2017;

1 square - Hansis et al. 2015). Blue cross: the mean from DGVMs with the box showing the 1 standard
 2 deviation range. Green triangles: downloaded from FAOSTAT website; the open triangle is primarily
 3 forest-related emissions, while the closed triangle includes emission from peat fires and peat drainage.
 4 Yellow inverted triangle: Greenhouse Gas Inventories (GHGI) LULUCF flux (Land Use Land Use
 5 Change and Forestry – part of AFOLU) based on country reports to UNFCCC (Grassi et al. 2018) –
 6 data for developing countries is only shown for 2006-2015 because reporting in many developing
 7 countries became more consistent/reliable after 2005. For more details on methods see Box 2.2:.

9 2.3.1.2.4 Processes responsible for the land sink

10 Just over half of total net anthropogenic CO₂ emissions (AFOLU and fossil fuels) were taken up by oceanic
 11 and land sinks (Table 2.3) (*robust evidence, high agreement*). The land sink was referred to in AR5 as the
 12 “residual terrestrial flux,” as it was not estimated directly, but calculated by difference from the other directly
 13 estimated fluxes in the budget (Table 2.3). In the 2018 budget (Le Quéré et al. 2018), the land sink term was
 14 instead estimated directly by DGVMs, leaving a budget imbalance of 2.2 GtCO₂ yr⁻¹ (sources overestimated
 15 or sinks underestimated). The budget imbalance may result from variations in oceanic uptake, or from
 16 uncertainties in fossil fuel or AFOLU emissions, as well as from land processes not included in DGVMs.

17
 18 The land sink is thought to be driven largely by the indirect effects of environmental change (e.g., climate
 19 change, increased atmospheric CO₂ concentration nitrogen deposition) on unmanaged and managed lands
 20 (*robust evidence, high agreement*). The land sink has generally increased since 1900 and was a net sink of
 21 11.7 ± 3.7 GtCO₂ yr⁻¹ during the period 2008 to 2017 (Table 2.3), absorbing 29% of global anthropogenic
 22 emissions of CO₂. The land sink has slowed the rise in global land-surface air temperature by 0.09 ± 0.02°C
 23 since 1982 (*medium confidence*) (Zeng et al. 2017).

24
 25 The rate of CO₂ removal by land accelerated from -0.026 ± 0.24 GtCO₂ yr⁻¹ during the warming period (1982
 26 to 1998) to -0.436 ± 0.260 GtCO₂ yr⁻¹ during the warming hiatus (1998-2012). One explanation is that
 27 respiration rates were lower during the warming hiatus (Ballantyne et al. 2017). However, the lower rate of
 28 growth in atmospheric CO₂ during the warming hiatus may have resulted, not from lower rates of respiration,
 29 but from declining emissions from AFOLU (lower rates of tropical deforestation and increased forest growth
 30 in northern mid-latitudes (Piao et al. 2018)). Changes in the growth rate of atmospheric CO₂, by themselves,
 31 do not identify the processes responsible, and the cause of the variation is uncertain.

32
 33 While year-to-year variability in the indirect land sink is high in response to climate variability, DGVM
 34 fluxes are far more influenced on decadal time scales by CO₂ fertilisation. A DGVM intercomparison (Sitch
 35 et al. 2015) for 1990 to 2009 found that CO₂ fertilisation alone contributed a mean global removal of -10.54
 36 ± 3.68 GtCO₂ yr⁻¹ (trend -0.444 ± 0.202 GtCO₂ yr⁻¹). Data from forest inventories around the world
 37 corroborate the modelled land sink (Pan et al. 2011a). The geographic distribution of the non-AFOLU land
 38 sink is less certain. While it seems to be distributed globally, its distribution between the tropics and non-
 39 tropics is estimated to be between 1:1 (Pan et al. 2011a) and 1:2 (Houghton et al. 2018).

40
 41 As described in Box 2.3, rising CO₂ concentrations have a fertilising effect on land, while climate has mixed
 42 effects; e.g., rising temperature increases respiration rates and may enhance or reduce photosynthesis
 43 depending on location and season, while longer growing seasons might allow for higher carbon uptake.
 44 However, these processes are not included in DGVMs, which may account for at least some of the land sink.
 45 For example, a decline in the global area burned by fires each year (Andela et al. 2017) accounts for an
 46 estimated net sink (and/or reduced emissions) of 0.5 GtCO₂ yr⁻¹ (Arora and Melton 2018) (*limited evidence,*
 47 *medium agreement*) (boreal forests represent an exception to this decline (Kelly et al. 2013)). The reduction
 48 in burning not only reduces emissions; it also allows more growth of recovering forests. There is also an
 49 estimated net carbon sink of about the same magnitude (0.5 GtCO₂ yr⁻¹) as a result of soil erosion from
 50 agricultural lands and redeposition in anaerobic environments where respiration is reduced (Wang et al.
 51 2017d) (*limited evidence, low agreement*). A recent study attributes an increase in land carbon to a longer-
 52 term (1860-2005) aerosol-induced cooling (Zhang et al. 2019). Recent evidence also suggests that DGVMs
 53 and Earth System Models underestimate the effects of drought on CO₂ emissions (Humphrey et al. 2018;
 54 Green et al. 2019; Kolus et al. 2019).

55 2.3.1.3 Gross emissions and removals contributing to AFOLU emissions

56 The modelled AFOLU flux of 5.5 ± 3.7 GtCO₂ yr⁻¹ over the period 2008 to 2017 represents a net value. It

1 consists of both gross emissions of CO₂ from deforestation, forest degradation, and the oxidation of wood
2 products, as well as gross removals of CO₂ in forests and soils recovering from harvests and agricultural
3 abandonment (Figure 2.4). The uncertainty of these gross fluxes is high because few studies report gross
4 fluxes from AFOLU. Houghton and Nassikas (2017) estimated gross emissions to be as high as 20.2 GtCO₂
5 yr⁻¹ (*limited evidence, low agreement*) (Figure 2.4), and even this may be an underestimate because the land-
6 use change data used from FAOSTAT (Tubiello et al. 2013) is itself a net of all changes within a country.
7

8 Gross emissions and removals of CO₂ result from rotational uses of land, such as wood harvest and shifting
9 cultivation, including regrowth. These gross fluxes are more informative for assessing the timing and
10 potential for mitigation than estimates of net fluxes, because the gross fluxes include a more complete
11 accounting of individual activities. Gross emissions from rotational land use in the tropics are approximately
12 37% of total CO₂ emissions, rather than 14%, as suggested by net AFOLU emissions (Houghton and
13 Nassikas 2018). Further, if the forest is replanted or allowed to regrow, gross removals of nearly the same
14 magnitude would be expected to continue for decades.
15

16 **2.3.1.4 Gross emissions and removals contributing to the non-anthropogenic land sink**

17 The *net* land sink averaged 11.7 GtCO₂ yr⁻¹ over 2008-2017 (*robust evidence, medium agreement*) (Table
18 2.3.2), but its gross components have not been estimated at the global level. There are many studies that
19 suggest increasing emissions of carbon due to indirect environmental effects and natural disturbance, for
20 example temperature-induced increases in respiration rates (Bond-Lamberty et al. 2018); increased tree
21 mortality (Brienen et al. 2015; Berdanier and Clark 2016; McDowell et al. 2018); and thawing permafrost
22 (Schuur et al. 2015). The global carbon budget indicates that land and ocean sinks have *increased* over the
23 last six decades in proportion to total CO₂ emissions (Le Quéré et al. 2018) (*robust evidence, high*
24 *agreement*). That means that any emissions must have been balanced by even larger removals (likely driven
25 by CO₂ fertilisation, climate change, nitrogen deposition, erosion and redeposition of soil carbon, a reduction
26 in areas burned, aerosol-induced cooling, and changes in natural disturbances, Box 2.3)
27

28 Climate change is expected to impact terrestrial biogeochemical cycles via an array of complex feedback
29 mechanisms that will act to either enhance or decrease future CO₂ emissions from land. Because the gross
30 emissions and removals from environmental changes are not constrained at present, the balance of future
31 positive and negative feedbacks remains uncertain. Estimates from climate models included in AR5, CMIP5
32 (Coupled Model Intercomparison Project, 5), exhibit large differences for different carbon and nitrogen cycle
33 feedbacks and how they change in a warming climate (Anav et al. 2013; Friedlingstein et al. 2006;
34 Friedlingstein, et al. 2014). The differences are in large part due to the uncertainty regarding how primary
35 productivity and soil respiration will respond to environmental changes, with many of the models not even
36 agreeing on the sign of change. Furthermore, many models do not include a nitrogen cycle, which may limit
37 the CO₂ fertilisation effect in the future (see Box 2.3). There is an increasing amount of observational data
38 available and methods to constrain models (e.g. Cox et al. 2013; Prentice, et al., 2015) which can reduce
39 uncertainty.
40

41 **2.3.1.5 Potential impact of mitigation on atmospheric CO₂ concentrations**

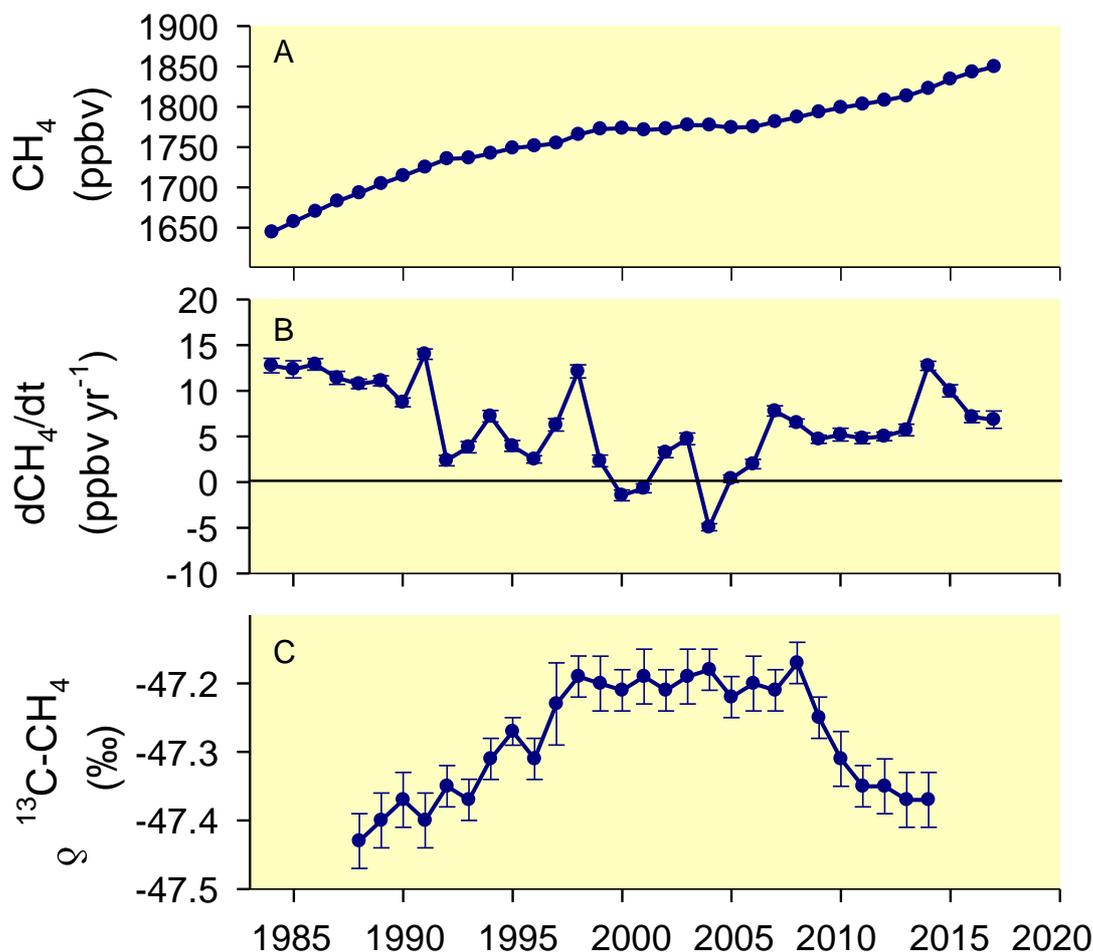
42 If CO₂ concentrations decline in the future as a result of low emissions and large negative emissions, the
43 global land and ocean sinks are expected to weaken (or even reverse). The oceans are expected to release
44 CO₂ back to the atmosphere when the concentration declines (Ciais et al. 2013a; Jones et al. 2016). This
45 means that to maintain atmospheric CO₂ and temperature at low levels, both the excess CO₂ from the
46 atmosphere and the CO₂ progressively outgassed from the ocean and land sinks will need to be removed.
47 This outgassing from the land and ocean sinks is called the “rebound effect” of the global carbon cycle (Ciais
48 et al. 2013a). It will reduce the effectiveness of negative emissions and increase the deployment level needed
49 to achieve a climate stabilisation target (Jackson et al. 2017; Jones et al. 2016) (*limited evidence, high*
50 *agreement*).
51

52 **2.3.2 Methane**

54 **2.3.2.1 Atmospheric trends**

55 In 2017, the globally averaged atmospheric concentration of CH₄ was 1850 ± 1 ppbv (Figure 2.8A).
56 Systematic measurements of atmospheric CH₄ concentrations began in the mid-1980s and trends show a

1 steady increase between the mid-1980s and early-1990s, slower growth thereafter until 1999, a period of no
 2 growth between 1999 and 2006, followed by a resumption of growth in 2007. The growth rates show very
 3 high inter-annual variability with a negative trend from the beginning of the measurement period until about
 4 2006, followed by a rapid recovery and continued high inter-annual variability through 2017 (Figure 2.8B).
 5 The growth rate has been higher over the past 4 years (*high confidence*) (Nisbet et al. 2019). The trend in
 6 $\delta^{13}\text{C-CH}_4$ prior to 2000 with less depleted ratios indicated that the increase in atmospheric concentrations
 7 was due to thermogenic (fossil) CH_4 emissions; the reversal of this trend after 2007 indicates a shift to
 8 biogenic sources (Figure 2.8C).



9
10
11 **Figure 2.8 Globally averaged atmospheric CH_4 mixing ratios (Frame A) and instantaneous rates of change**
 12 **(Frame B) and C isotope /variation (Frame C) Data sources: NOAA/ESRL**
 13 **(www.esrl.noaa.gov/gmd/ccgg/trends_ch4/)(Dlugokencky et al. 1994) and Schaefer et al. (2016).**
 14

15 Understanding the underlying causes of temporal variation in atmospheric CH_4 concentrations is an active
 16 area of research. Several studies concluded that inter-annual variability of CH_4 growth was driven by
 17 variations in natural emissions from wetlands (Rice et al. 2016; Bousquet et al. 2006; Bousquet et al. 2011;
 18 Bousquet et al. 2011b). These modelling efforts concluded that tropical wetlands were responsible for
 19 between 50 and 100% of the inter-annual fluctuations and the renewed growth in atmospheric concentrations
 20 after 2007. However, results were inconsistent for the magnitude and geographic distribution of the wetland
 21 sources between the models. Pison et al. (2013) used two atmospheric inversion models and the ORCHIDEE
 22 model and found greater uncertainty in the role of wetlands in inter-annual variability between 1990 and
 23 2009 and during the 1999-2006 pause. Poulter et al. (2017) used several of biogeochemical models and
 24 inventory-based wetland area data to show that wetland CH_4 emissions increases in the boreal zone have
 25 been offset by decreases in the tropics and concluded that wetlands have not contributed significantly to
 26 renewed atmospheric CH_4 growth.
 27

1 The models cited above assumed that atmospheric hydroxyl radical (OH) sink over the period analysed did
2 not vary. OH reacts with CH₄ as the first step toward oxidation to CO₂. In global CH₄ budgets, the
3 atmospheric OH sink has been difficult to quantify because its short lifetime (~1 second) and its distribution
4 is controlled by precursor species that have non-linear interactions (Taraborrelli et al., 2012; Prather et al.,
5 2017). Understanding of the atmospheric OH sink has evolved recently. The development of credible time
6 series of methyl chloroform (MCF: CH₃CCl₃) observations offered a way understand temporal dynamics of
7 OH abundance and applying this to global budgets further weakened the argument for the role of wetlands in
8 determining temporal trends since 1990. Several authors used the MCF approach and concluded that changes
9 in the atmospheric OH sink explained a large portion of the suppression in global CH₄ concentrations
10 relative to the pre-1999 trend (Turner et al. 2017; Rigby et al. 2013; McNorton et al. 2016). These studies
11 could not reject the null hypothesis that OH has remained constant in recent decades and they did not suggest
12 a mechanism for the inferred OH concentration changes (Nisbet et al. 2019). Nicely et al. (2018) used a
13 mechanistic approach and demonstrated that variation in atmospheric OH was much lower than what MCF
14 studies and found that positive trends in OH due to the effects of water vapour, nitrogen oxides (NO_x),
15 tropospheric ozone, and expansion of the tropical Hadley cells offsets the decrease in OH that is expected
16 from increasing atmospheric CH₄ concentrations.

17
18 The depletion of δ¹³C_{atm} beginning in 2009 could be due to changes in several sources. Decreased fire
19 emissions combined with increased tropical wetland emissions compared to earlier years could explain the
20 δ¹³C perturbations to atmospheric CH₄ sources (Worden et al. 2017; Schaefer et al. 2016). However, because
21 tropical wetland emissions are higher in the Southern Hemisphere, and the remote sensing observations show
22 that CH₄ emissions increases are largely in the north tropics (Bergamaschi et al. 2013; Melton et al. 2013;
23 Houweling et al. 2014), an increased wetland source does not fit well with the southern hemisphere δ¹³C
24 observations. New evidence shows that tropical wetland CH₄ emissions are significantly underestimated,
25 perhaps by a factor of 2, because estimates do not account for release by tree stems (Pangala et al. 2017).
26 Several authors have concluded that agriculture is a more probable source of increased emissions, and
27 particularly from rice and livestock in the tropics, which is consistent with inventory data (Wolf et al. 2017;
28 Patra et al. 2016; Schaefer et al. 2016).

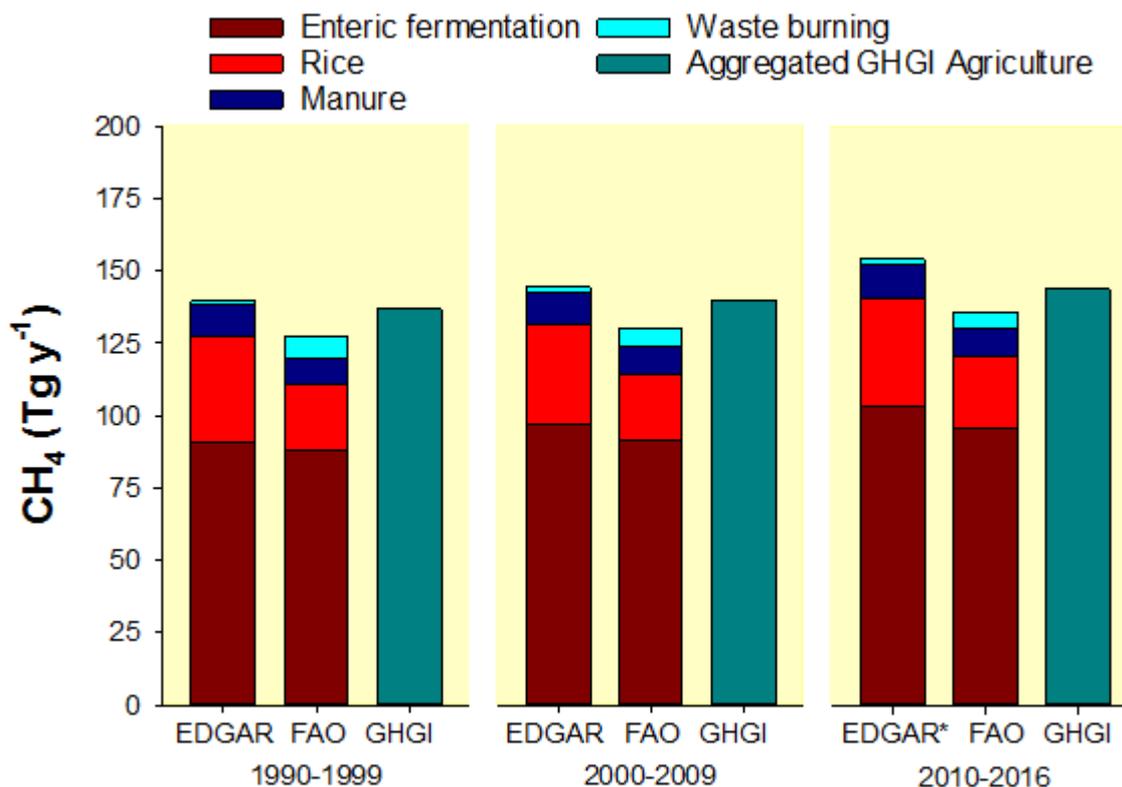
29
30 The importance of fugitive emissions in the global atmospheric accumulation rate is growing (*medium*
31 *evidence, high agreement*). The increased production of natural gas in the US from the mid 2000's is of
32 particular interest because it coincides with renewed atmospheric CH₄ growth (Rice et al. 2016; Hausmann et
33 al. 2015). Reconciling increased fugitive emissions with increased isotopic depletion of atmospheric CH₄
34 indicates that there are *likely* multiple changes in emissions and sinks that affect atmospheric accumulation
35 (*medium confidence*).

36
37 With respect to atmospheric CH₄ growth rates, we conclude that there is significant and ongoing
38 accumulation of CH₄ in the atmosphere (*very high confidence*). The reason for the pause in growth rates and
39 subsequent renewed growth is at least partially associated with land use and land use change. Evidence that
40 variation in the atmospheric OH sink plays a role in the year to year variation of the CH₄ is accumulating,
41 but results are contradictory (*medium evidence, low agreement*) and refining this evidence is constrained by
42 lack of long-term isotopic measurements at remote sites, particularly in the tropics. Fugitive emissions *likely*
43 contribute to the renewed growth after 2006 (*medium evidence, high agreement*). Additionally, the recent
44 depletion trend of ¹³C isotope in the atmosphere indicates that growth in biogenic sources explains part of the
45 current growth and that biogenic sources make up a larger proportion of the source mix compared to the
46 period before 1997 (*robust evidence, high agreement*). In agreement with the findings of AR5, we conclude
47 that wetlands are important drivers of inter-annual variability and current growth rates (*medium evidence,*
48 *high agreement*). Ruminants and the expansion of rice cultivation are also important contributors to the
49 current growth trend (*medium evidence, high agreement*).

50 51 **2.3.2.2 Land use effects**

52 Agricultural emissions are predominantly from enteric fermentation and rice, with manure management and
53 waste burning contributing small amounts (Figure 2.9). Since 2000, livestock production has been
54 responsible for 33% of total global emissions and 66% of agricultural emissions (Source: EDGAR 4.3.2
55 database, accessed May 2018, (USEPA 2012; Tubiello et al. 2014; Janssens-Maenhout et al. 2017b). Asia
56 has the largest livestock emissions (37%) and emissions in the region have been growing by around 2% per

1 year over the same period. North America is responsible for 26% and emissions are stable; Europe is
 2 responsible for around 8% of emissions, and these are decreasing slightly (<1% per year). Africa is
 3 responsible for 14%, but emissions are growing fastest in this region at around 2.5% y^{-1} . In Latin America
 4 and the Caribbean, livestock emissions are decreasing at around 1.6% per year and the region makes up 16%
 5 of emissions. Rice emissions are responsible for about 24% of agricultural emissions, and 89% of these are
 6 from Asia. Rice emissions are increasing by 0.9% per year in that region. These trends are predicted to
 7 continue through 2030 (USEPA 2013).



8 **Figure 2.9 Average agricultural CH₄ emissions estimates from 1990. Sub-sectorial agricultural emissions are**
 9 **based on the Emissions Database for Global Atmospheric Research (EDGAR v4.3.2; Janssens-**
 10 **Maenhout et al. 2017a); FAOSTAT (Tubiello et al. 2013); and National GHGI data (Grassi et al. 2018).**
 11 **GHGI data are aggregate values for the sector.**

12 * Note that EDGAR data are complete only through 2012; the data in the right-hand panel represent the
 13 three years 2010-2012 and are presented for comparison.
 14

15

16

17

18

19

20

21

22

23

24

25

26

27

28

29

30

31

32

Upland soils are a net sink of atmospheric CH₄, but soils both produce and consume the gas. On the global
 scale climatic zone, soil texture, and land cover have an important effect on CH₄ uptake in upland soils (Tate
 2015; Yu et al. 2017; Dutaur and Verchot 2007). Boreal soils take up less than temperate or tropical soils,
 coarse textured soils take up more CH₄ than medium and fine textured soils, and forests take up more than
 other ecosystems. Low levels of nitrogen fertilisation or atmospheric deposition can affect the soil microbial
 community and stimulate soil CH₄ uptake in nitrogen limited soils, while higher fertilisation rates decrease
 uptake (Edwards et al. 2018; Zhuang et al., 2013). Globally, N fertilisation on agricultural lands may have
 suppressed CH₄ oxidation by as much as 26 Tg between 1998 and 2004 (Zhuang et al., 2013)(*low confidence* ,
low agreement). The effect of N additions is cumulative and repeated fertilisation events have progressively
 greater suppression effects (*robust evidence, high agreement*) (Tate 2015). Other factors like higher
 temperatures, increased atmospheric concentrations, and changes in rainfall patterns stimulate soil CH₄
 consumption in unfertilised ecosystems. Several studies (Yu et al. 2017; Xu et al. 2016; Curry 2009) have
 shown that globally, uptake has been increasing during the second half of the 20th century and it is expected
 to continue to increase by as much as 1 Tg in the 21st century, particularly in forests and grasslands (*medium*
evidence, high agreement).

Northern peatlands (40°-70°N) are a significant source of atmospheric CH₄, emitting about 48 Tg CH₄, or

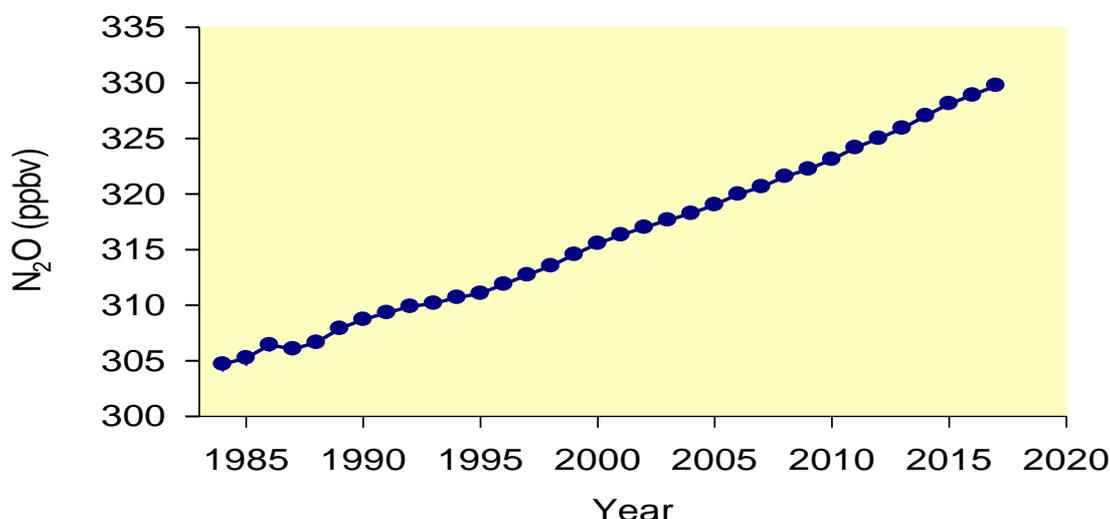
1 about 10% of the total emissions to the atmosphere (Zhuang et al. 2006; Wuebbles and Hayhoe 2002). CH₄
 2 emissions from natural northern peatlands are highly variable with the highest rate from fens (*medium*
 3 *evidence, high agreement*). Peatland management and restoration, alters the exchange of CH₄ with the
 4 atmosphere (*medium evidence, high agreement*). Management of peat soils typically converts them from CH₄
 5 sources to sinks (Augustin et al. 2011; Strack and Waddington 2008; Abdalla et al. 2016) (*robust evidence,*
 6 *high agreement*). While restoration decreases CO₂ emissions (see Section 4.9.4), CH₄ emissions often
 7 increase relative to the drained conditions (*robust evidence, high agreement*) (Osterloh et al. 2018; Christen
 8 et al. 2016; Koskinen et al. 2016; Tuittila et al. 2000; Vanselow-Algan et al. 2015; Abdalla et al. 2016).
 9 Drained peatlands are usually considered to be negligible methane sources, but they emit CH₄ under wet
 10 weather conditions and from drainage ditches (Drösler et al. 2013; Sirin et al. 2012). While ditches cover
 11 only a small percentage of the drained area, emissions can be sufficiently high that drained peatlands emit
 12 comparable CH₄ as undrained ones (*medium evidence, medium agreement*) (Sirin et al. 2012; Wilson et al.
 13 2016).

14
 15 Because of the large uncertainty in the tropical peatland area, estimates of the global flux are highly
 16 uncertain. A meta-analysis of the effect of conversion of primary forest to rice production showed that
 17 emissions increased by a factor of 4 (*limited evidence, high agreement*) (Hergoualc'h and Verchot, 2012).
 18 For land uses that required drainage, emissions decreased by a factor of 3 (*limited evidence, high*
 19 *agreement*). There are no representative measurements of emissions from drainage ditches in tropical
 20 peatlands.

21 2.3.3 Nitrous Oxide

22 2.3.3.1 Atmospheric trends

23
 24 The atmospheric abundance of N₂O has increased since 1750, from a pre-industrial concentration of 270
 25 ppbv to 330 ppbv in 2017 (U.S. National Oceanographic and Atmospheric Agency, Earth Systems Research
 26 Laboratory; Figure 2.10) (*high agreement, robust evidence*). The rate of increase has also increased, from
 27 approximately 0.15 ppbv yr⁻¹ 100 years ago, to 0.85 ppbv yr⁻¹ over the period 2001 to 2015 (Wells et al.
 28 2018). Atmospheric N₂O isotopic composition (^{14/15}N) was relatively constant during the pre-industrial
 29 period (Prokopiou et al. 2018) and shows a decrease in the δ¹⁵N as the N₂O mixing ratio in the atmosphere
 30 has increased between 1940 and 2005. This recent decrease indicates as that terrestrial sources are the
 31 primary driver of increasing trends and marine sources contribute around 25% (Snider et al. 2015).
 32 Microbial denitrification and nitrification processes are responsible for more than 80% of total global N₂O
 33 emissions, which includes natural soils, agriculture, and oceans, with the remainder coming from non-
 34 biological sources such as biomass burning and fossil-fuel combustion (Fowler et al. 2015). The isotopic
 35 trend also indicates a shift from denitrification to nitrification as the primary source of N₂O as a result of the
 36 use of synthetic nitrogen (N) fertiliser (*high evidence, high agreement*) (Park et al. 2012; Toyoda et al. 2013;
 37 Snider et al. 2015; Prokopiou et al. 2018).
 38

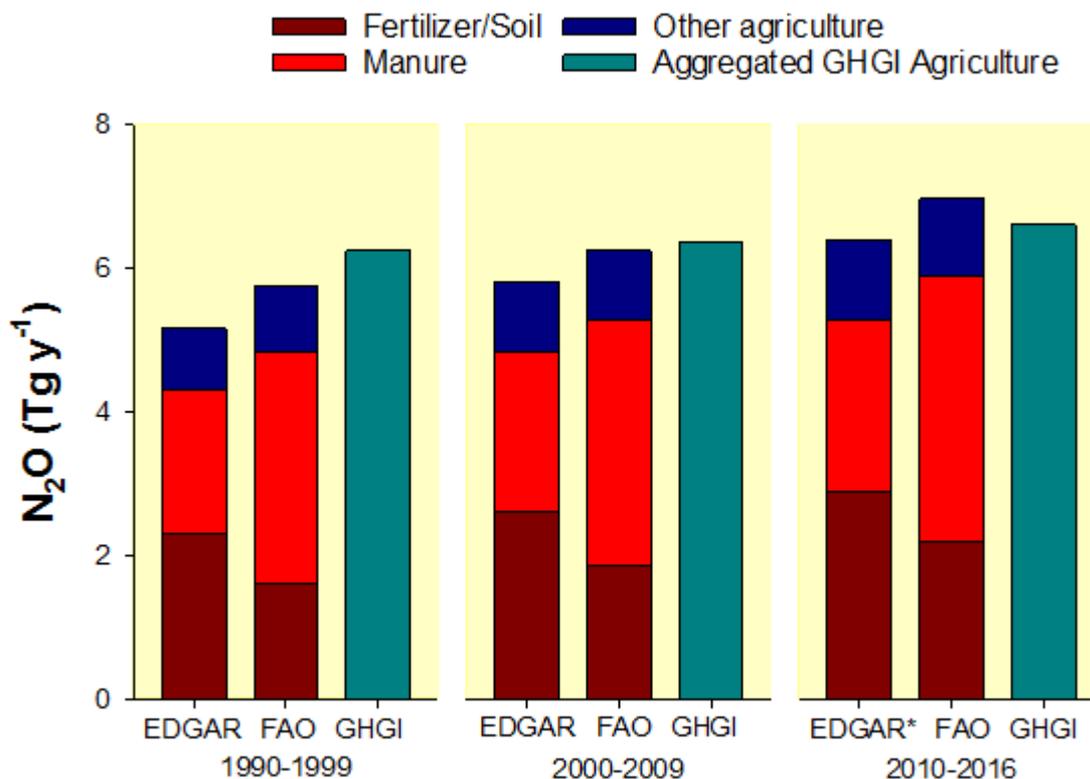


39
 40 **Figure 2.10 Globally averaged atmospheric N₂O mixing ratios since 1984. Data sources: NOAA/ESRL**

1 **Global Monitoring Division (<https://www.esrl.noaa.gov/gmd/hats/combined/N2O.html>).**

2
3 The three independent sources of N₂O emissions estimates from agriculture at global, regional, and national
4 levels are: U.S.E.P.A., EDGAR and FAOSTAT (USEPA 2013; Tubiello et al. 2015; Janssens-Maenhout et
5 al. 2017). EDGAR and FAOSTAT have temporal resolution beyond 2005 and we these databases compare
6 well with national inventory data (Figure 2.10). USEPA has historical estimates through 2005 and
7 projections thereafter. The independent data use IPCC methods, with Tier 1 emission factors and national
8 reporting of activity data. Tier 2 approaches are also available based on top-down and bottom-up
9 approaches. Recent estimates using inversion modelling and process models estimate total annual global
10 N₂O emissions of 16.1-18.7 (bottom-up) and 15.9-17.7 Tg N (top-down), demonstrating relatively close
11 agreement (Thompson et al. 2014). Agriculture is the largest source and has increased with the
12 extensification and intensification. Recent modelling estimates of terrestrial sources show a higher emissions
13 range that is slightly more constrained than what was reported in AR5: approximately 9 (7–11) Tg N₂O-N yr⁻¹
14 (Saikawa et al. 2014; Tian et al. 2016) compared to 6.6 (3.3–9.0) Tg N₂O-N yr⁻¹ (Ciais et al. 2013a).
15 Estimates of marine N₂O emissions are between 2.5 and 4.6 Tg N₂O-N yr⁻¹; (Buitenhuis et al., 2017;
16 Saikawa et al., 2014).

17
18 To conclude, N₂O is continuing to accumulate in the atmosphere at an increasingly higher rate (*very high*
19 *confidence*), driven primarily by increases in manure production and synthetic N fertiliser use from the mid-
20 20th century onwards (*high confidence*). Findings since AR5 have constrained regional and global estimates
21 of annual N₂O emissions and improved our understanding of the spatio-temporal dynamics of N₂O
22 emissions, with soil rewetting and freeze-thaw cycles, which important determinants of total annual emission
23 fluxes in some regions (*medium confidence*).



24
25 **Figure 2.11 Average agricultural N₂O emissions estimates from 1990. Sub-sectorial agricultural emissions**
26 **are based on the Emissions Database for Global Atmospheric Research (EDGAR v4.3.2; Janssens-**
27 **Maenhout et al. 2017a); FAOSTAT (Tubiello et al. 2013); and National GHGI data (Grassi et al. 2018).**
28 **GHGI data are aggregate values for the sector.**

29 * Note that EDGAR data are complete only through 2012; the EDGAR data in the right-hand panel
30 represent the three years 2010-2012 and are presented for comparison.

31 32 2.3.3.2 Land use effects

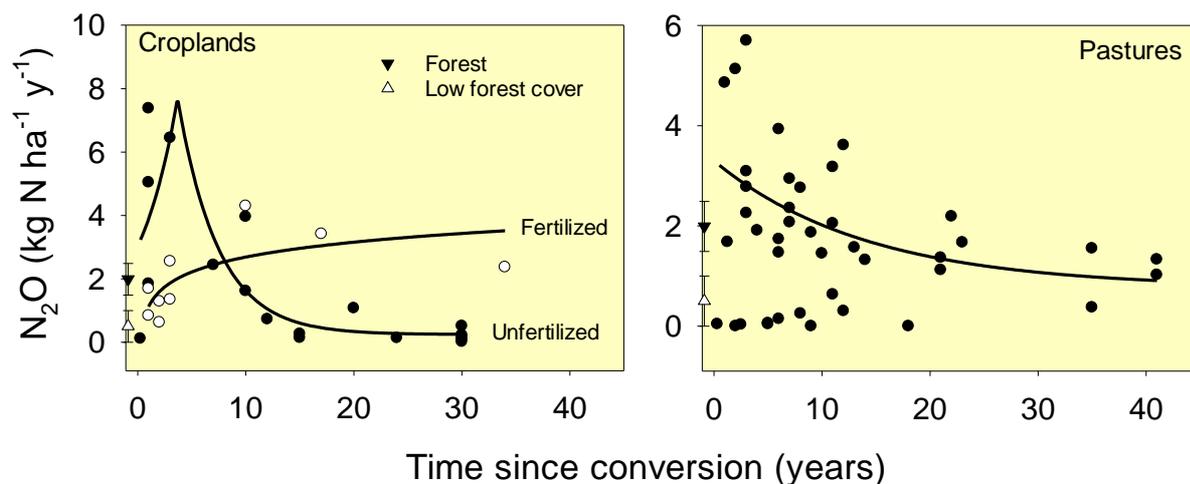
33 Agriculture is responsible for approximately two-thirds of N₂O emissions (*robust evidence, high agreement*)

1 (Janssens-Maenhout et al. 2017). Total emissions from this sector are the sum of direct and indirect
2 emissions. Direct emissions from soils are the result of mineral fertiliser and manure application, manure
3 management, deposition of crop residues, cultivation of organic soils and inorganic N inputs through
4 biological nitrogen fixation. Indirect emissions come from increased warming, enrichment of downstream
5 water bodies from runoff, and downwind N deposition on soils. The main driver of N₂O emissions in
6 croplands is a lack of synchronisation between crop N demand and soil N supply, with approximately 50%
7 of N applied to agricultural land not taken up by the crop (Zhang et al. 2017). Cropland soils emit over 3 Tg
8 N₂O-N yr⁻¹ (*medium evidence, high agreement*) (Janssens-Maenhout et al. 2017; Saikawa et al. 2014).
9 Regional inverse modelling studies show larger tropical emissions than the inventory approaches and they
10 show increases in N₂O emissions from the agricultural sector in South Asia, Central America, and South
11 America (Saikawa et al. 2014; Wells et al. 2018).

12
13 Emissions of N₂O from pasturelands and rangelands have increased by as much as 80% since 1960 due to
14 increased manure production and deposition (*robust evidence, high agreement*) (de Klein et al. 2014; Tian et
15 al. 2018; Chadwick et al. 2018; Dangal et al. 2019; Cardenas et al. 2019). Studies consistently report that
16 pasturelands and rangelands are responsible for around half of the total agricultural N₂O emissions
17 (Davidson 2009; Oenema et al. 2014; Dangal et al. 2019). An analysis by Dangal et al. (2019) shows that
18 while managed pastures make up around one-quarter of the global grazing lands, they contribute 86% of the
19 net global N₂O emissions from grasslands and that more than half of these emissions are related to direct
20 deposition of livestock excreta on soils.

21
22 Many studies calculate N₂O emissions from a linear relationship between nitrogen application rates and N₂O
23 emissions. New studies are increasingly finding nonlinear relationships, which means that N₂O emissions per
24 hectare are lower than the Tier 1 EFs (IPCC 2003) at low nitrogen application rates, and higher at high
25 nitrogen application rates (*robust evidence, high agreement*) (Shcherbak et al. 2014; van Lent et al. 2015;
26 Satria 2017). This not only has implications for how agricultural N₂O emissions are estimated in national and
27 regional inventories, which now often use a linear relationship between nitrogen applied and N₂O emissions,
28 it also means that in regions of the world where low nitrogen application rates dominate, increases in
29 nitrogen fertiliser use would generate relatively small increases in agricultural N₂O emissions. Decreases in
30 application rates in regions where application rates are high and exceed crop demand for parts of the growing
31 season are likely to have very large effects on emissions reductions (*medium evidence, high agreement*).

32
33 Deforestation and other forms of land-use change alter soil N₂O emissions. Typically, N₂O emissions
34 increase following conversion of native forests and grasslands to pastures or croplands (McDaniel et al.
35 2019; van Lent et al. 2015). This increase lasts from a few years to a decade or more, but there is a trend
36 toward decreased N₂O emissions with time following land use change and ultimately lower N₂O emissions
37 than had been occurring under native vegetation, in the absence of fertilisation (**Figure 2.12**) (Meurer et al.
38 2016; van Lent et al. 2015) (*medium evidence, high agreement*). Conversion of native vegetation to fertilised
39 systems typically leads to increased N₂O emissions over time, with the rate of emission often being a
40 function of nitrogen fertilisation rates, but this response can be moderated by soil characteristics and water
41 availability (*medium evidence, high agreement*) (van Lent et al. 2015; Meurer et al. 2016). Restoration of
42 agroecosystems to natural vegetation, over the period of one to two decades does not lead to recovery of N₂O
43 emissions to the levels of the original vegetation (McDaniel et al. 2019). To conclude, findings since AR5
44 increasingly highlight the limits of linear N₂O emission factors, particularly from field to regional scales,
45 with emissions rising nonlinearly at high nitrogen application rates (*high confidence*). Emissions from
46 unfertilised systems often increase and then decline over time with typically lower emissions than was the
47 case under native vegetation (*high confidence*).



1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36
37
38
39
40

Figure 2.12 Effect of time since conversion on N_2O fluxes in unfertilised (black circles) and fertilised (white circles) tropical croplands (left frame) and in unfertilised tropical pastures (right frame). Average N_2O flux and 95% confidence intervals are given for upland forests (black inverted triangle) and low canopy forests (white inverted triangle), for comparison. The solid lines represent the trends for unfertilised and fertilised cases. Data source: (van Lent et al. 2015).

While soil emissions are the predominant source in agriculture, other sources are important or their importance is only just emerging. Biomass burning is responsible for approximately $0.7 \text{ Tg N}_2\text{O-N yr}^{-1}$ ($0.5\text{--}1.7 \text{ Tg N}_2\text{O-N yr}^{-1}$) or 11% of total gross anthropogenic emissions due to the release of N_2O from the oxidation of organic nitrogen in biomass (van der Werf et al. 2013). This source includes crop residue burning, forest fires, household cook stoves, and prescribed savannah, pasture and cropland burning. Aquaculture is currently not accounted for in most assessments or compilations. While it is currently responsible for less than $0.1 \text{ Tg N}_2\text{O-N yr}^{-1}$, is one of the fastest growing sources of anthropogenic N_2O emissions (Williams and Crutzen 2010; Bouwman et al. 2013) (*limited evidence, high agreement*). Finally, increased nitrogen deposition from terrestrial sources is leading to greater indirect N_2O emissions, particularly since 1980 (*moderate evidence, high agreement*). (Tian et al. 2018, 2016). In marine systems, deposition is estimated to have increased the oceanic N_2O source by $0.2 \text{ Tg N}_2\text{O-N yr}^{-1}$ or 3% of total gross anthropogenic emissions (Suntharalingam et al. 2012).

Box 2.2: Methodologies for estimating national to global scale anthropogenic land carbon fluxes

Bookkeeping/accounting models (Houghton et al. 2012b; Hansis et al. 2015; Houghton and Nassikas 2017) calculate changes in biomass and soils that result from changes in land activity using data on biomass density and rates of growth/decomposition, typically from ground-based inventory data collection (field measurements of carbon in trees and soils). The approach includes only those changes directly caused by major categories of land-use change and management. The models do not explicitly include the indirect effects to changing environmental conditions, although some effects are implicit in biomass, growth rates and decay rates used. Thus, the models may overestimate past fluxes. The bookkeeping models include fluxes from peatland burning based on GFED estimates (Global Fire Emissions Database, (Randerson et al. 2015).)

Dynamic Global Vegetation Models (DGVMs) simulate ecological processes, such as photosynthesis, respiration, allocation, growth, decomposition etc., driven by environmental conditions (climate variability, climate change, CO_2 , nitrogen concentrations). Models vary with respect to the processes included, with many since AR5 now including forest management, fire, N, and other management (Sitch et al. 2005; Le Quéré et al. 2018). Models are forced with increasing atmospheric CO_2 and changing climate, and run with and without “land use change” (land cover and forest harvest) to differentiate the anthropogenic effects from

1 the indirect effects of climate and CO₂ - the “land sink”. Thus, indirect effects are explicitly included. This
2 approach also includes a “lost atmospheric sink capacity”, or the carbon uptake due to environmental effects
3 on forests that does not happen once the forests are removed (Pongratz et al. 2010).
4

5 **Integrated Assessment Models (IAMs)** use story-lines to construct alternative future scenarios of GHG
6 emissions and atmospheric concentrations within a global socio-economic framework, including projections
7 of AFOLU based on assumptions of, for example, crop yields, population growth, bioenergy use (See Cross-
8 Chapter Box 1: Scenarios, Chapter 1). Some models include simplified DGVMs, which may include climate
9 and CO₂ effects, while others use AFOLU emissions from other sources.
10

11 **Earth system models (ESMs)** couple DGVMs, surface hydrology and energy exchange models with a
12 atmosphere, ocean, and sea ice models, enabling exploration of feedbacks between climate change and the
13 carbon cycle (e.g., warming effects increase soil and plant respiration and lead to higher atmospheric CO₂
14 concentrations, which in turn promote plant growth) (Friedlingstein et al. 2014). They sometimes include
15 numerical experiments with and without land-use change to diagnose the anthropogenic AFOLU flux
16 (Lawrence et al. 2016).
17

18 **Satellite data** can be used as a proxy for plant activity (e.g. greenness) and to map land cover, vegetation
19 fires and biomass density. Algorithms, models and independent data are used to calculate fluxes of CO₂ from
20 satellite data, although calculating the net carbon flux is difficult because of the lack of information on the
21 respiratory flux. Some active satellite sensors (LiDAR) are able to measure three-dimensional structure in
22 woody vegetation, which is closely related to biomass density (Zarin et al. 2016a; Baccini et al. 2012;
23 Saatchi et al. 2011). Together with land-cover change data, these estimates of biomass density can be used to
24 provide observational-based estimates of fluxes due to changes in forest area (e.g., (Tyukavina et al. 2015;
25 Harris et al. 2015; Baccini et al. 2012) or degradation (Baccini et al. 2017). Satellite estimates of biomass
26 vary considerably (Mitchard et al. 2013; Saatchi et al. 2015; Avitabile et al. 2016); data are available only for
27 recent decades; methods generally assume that all losses of carbon are immediately released to the
28 atmosphere; and changes in soil carbon are generally ignored. The approach implicitly includes indirect and
29 natural disturbance effects as well as direct anthropogenic effects.
30

31 **Atmospheric Inversions** use observations of atmospheric concentrations with a model of atmospheric
32 transport, based on data for wind speed and direction, to calculate implied emissions (Gatti et al. 2014; Liu et
33 al. 2017a; van der Laan-Luijkx et al. 2017). Since AR5 there has been an increase in availability of
34 concentration data from flux tower networks and satellites, enabling better global coverage at finer spatial
35 scales and some national estimates (e.g. in the UK inverse techniques are used together with national GHG
36 inventories). A combination of concentrations of different gases and isotopes enables the separation of fossil,
37 ocean and land fluxes. However, inversions give only the net flux of CO₂ from land; they cannot separate
38 natural and anthropogenic fluxes.
39

40 **Micrometeorological flux measurements:** Data on CO₂ concentrations and air movements recorded on
41 instrumented towers enable calculation of CO₂ flux at the ecosystem scale. Global and regional Flux
42 Networks (FluxNet (Global), AsiaFlux, Ameriflux (North America), ICOS (EU), NEON (USA), and others)
43 contribute to a global flux data base, which is used to verify the results of modelling, inventory and remote
44 sensing studies.
45

46 **FAOSTAT:** The United Nations Food and Agricultural Organization has produced country level estimates
47 of greenhouse gas emissions (Tubiello et al. 2013) from agriculture (1961–2016) and land use (1990–2016)
48 using a globally consistent methodological approach based largely on IPCC Tier 1 methods of the 2006
49 IPCC Guidelines (FAO 2015b). FAO emissions estimates were used as one of the three database inputs into
50 the AR5 WGIII AFOLU chapter. Non-CO₂ emissions from agriculture are estimated directly from national
51 statistics of activity data reported by countries to FAO. CO₂ emissions from land use and land-use change are
52 computed mostly at Tier 1, albeit at fine geospatial scales to capture effects from peatland degradation and
53 biomass fires (Rossi et al. 2016). Emissions from forest land and deforestation are based on the IPCC carbon
54 stock change method, thus constituting a Tier 3 estimate relying on country statistics of carbon stocks and
55 forest area collected through the FAO FRA. The carbon flux is estimated assuming instantaneous emissions
56 in the year of forest area loss, and changes in carbon stocks within extant forests, but does not distinguish

1 “managed” and “unmanaged” forest areas, albeit it treats separately emissions from primary, secondary and
 2 planted forest (Federici et al. 2015).
 3

4 **Country Reporting of GHG Inventories (GHGIs):** All Parties to the UNFCCC are required to report
 5 national GHG Inventories (GHGIs) of anthropogenic emissions and removals. Reporting requirements are
 6 differentiated between developed and developing countries. Because of the difficulty of separating direct
 7 anthropogenic fluxes from indirect or natural fluxes, the (IPCC 2003) adopted the “managed land” concept
 8 as a proxy to facilitate GHGI reporting. All GHG fluxes on “managed land” are defined as anthropogenic,
 9 with each country applying their own definition of “managed land” (i.e. “where human interventions and
 10 practices have been applied to perform production, ecological or social functions” (IPCC 2006)). Fluxes may
 11 be determined on the basis of changes in carbon stocks (e.g., from forest inventories) or by activity data (e.g.
 12 area of land cover change management activity multiplied by emission factors or with modelled fluxes).
 13 Depending on the specific methods used, GHGIs include all direct anthropogenic effects and may include the
 14 indirect anthropogenic effects of environmental change (generally sinks) and natural effects (see Section
 15 2.3.1.2). GHG fluxes from “unmanaged land” are not reported in GHGIs because they are assumed to be
 16 non-anthropogenic. The reported estimates may then be filtered through agreed “accounting rules” - i.e. what
 17 countries actually count towards their mitigation targets (Cowie AL et al. 2007; Lee, D. and Sanz 2017). The
 18 accounting aims to better quantify the additional mitigation actions by, for example, factoring out the impact
 19 of natural disturbances and forest age-related dynamics (Canadell et al. 2007; Grassi et al. 2018).
 20

25 **Box 2.3: CO₂ fertilisation and enhanced terrestrial uptake of carbon**

26
 27 All Dynamic Global Vegetation Models (DGVMs) and Earth System Models (ESMs) represent the CO₂
 28 fertilisation effect ((Le Quéré et al. 2017; Hoffman et al. 2014). There is *high confidence* that elevated CO₂
 29 results in increased short-term CO₂ uptake per unit leaf area (Swann et al. 2016; Field et al. 1995; Donohue
 30 et al. 2013);. However, whether this increased CO₂ uptake at the leaf level translates into increased growth
 31 for the whole plant differs among plant species and environments because growth is constrained by whole-
 32 plant resource allocation, nutrient limitation (e.g., nitrogen (N), phosphorus (P), potassium (K) and soil water
 33 and light limitations (Körner 2006; Peñuelas et al. 2017; Friend et al. 2014a). Interactions between plants and
 34 soil microbes, further modulate the degree of nutrient limitation on CO₂ fertilisation (Terrer et al. 2017).
 35

36 At the ecosystems level, enhanced CO₂ uptake at decadal or longer time scales depends on changes in plant
 37 community composition and ecosystem respiration, as well disturbance and natural plant mortality (De
 38 Kauwe et al., 2016; Farrior et al., 2015; Keenan et al., 2017; Sulman et al, 2019). The results of FACE
 39 experiments (free-air carbon dioxide enrichment) over two decades are highly variable because of these
 40 factors (Norby et al. 2010; Körner 2015; Feng et al. 2015; Paschalis et al. 2017; Terrer et al. 2017; Du et al.
 41 2019b). Under higher atmospheric CO₂ concentrations, the ratio of CO₂ uptake to water loss (water use
 42 efficiency, WUE), increases and enhances drought tolerance of plants (*high confidence*) (Berry et al., 2010;
 43 Ainsworth & Rogers, 2007).
 44

45 Long-term CO₂ and water vapour flux measurements show that WUE in temperate and boreal forests of the
 46 Northern Hemisphere has increased more than predicted by photosynthetic theory and models over the past
 47 two decades (*high confidence*) (Keenan et al. 2013; Laguë and Swann 2016b). New theories have emerged
 48 on how CO₂ uptake by trees is related to water loss and to the risk of damaging xylem (water conducting
 49 tissues) in the trunk and branches (Wolf et al. 2016a; Anderegg et al. 2018a). Tree ring studies of stable
 50 carbon and oxygen isotopes also detected increased WUE in recent decades (Battipaglia et al. 2013; Silva
 51 and Anand 2013; van der Sleen et al. 2014). Yet, tree ring studies often fail to show acceleration of tree
 52 growth rates in support of CO₂ fertilisation, even when they show increased WUE (van der Sleen et al.
 53 2014). The International Tree Ring Data Bank (ITRDB) indicated that only about 20% of the sites in the
 54 database showed increasing trends in tree growth that cannot be explained by climate variability, nitrogen
 55 deposition, elevation, or latitude. Thus there is *limited evidence (low agreement)* among observations of
 56 enhanced tree growth due to CO₂ fertilisation of forests during the 20th century (Gedalof and Berg 2010).

1
2 In grasslands, although it is possible for CO₂ fertilisation to alleviate the impacts of drought and heat stress
3 on net carbon uptake (Roy et al. 2016b), there is *low confidence* about its projected magnitude. Because of
4 its effect on water use efficiency, CO₂ fertilisation is expected to be pronounced in semi-arid habitats; and
5 because of different metabolic pathways, C₃ plants are expected to be more sensitive to elevated CO₂
6 concentrations than C₄ grasses (Donohue et al. 2013; Morgan et al. 2011; Derner et al. 2003). Neither of
7 these expectations was observed over a 12-year study of elevated CO₂ in a grassland system: enhanced
8 growth was not observed during dry summers, and growth of C₄ grasses was unexpectedly stimulated, while
9 growth of C₃ grasses was not (Reich et al. 2014, 2018).

10
11 There is *medium confidence* that CO₂ fertilisation effects have increased water use efficiency in crops and
12 thus reduced agricultural water use per unit of crop produced (Deryng et al. 2016b; Nazemi and Wheeler
13 2015; Elliott et al. 2014). This effect could lead to near-term continued greening of agricultural areas.
14 However, current assessments of these effects are based on limited observations, mostly from the temperate
15 zone (Deryng et al. 2016a).

16
17 One line of evidence for CO₂ fertilisation is the increasing land sink (“the residual land sink” in AR5) over
18 the last 50 years as the atmospheric CO₂ concentration has increased (Los 2013; Sitch et al. 2015b; Campbell
19 et al. 2017; Keenan and Riley 2018). A combined analysis of atmospheric inverse analyses, ecosystem
20 models, and forest inventory data concluded that 60% of the recent terrestrial carbon sink can be directly
21 attributed to increasing atmospheric CO₂ (Schimel et al. 2015). A global analysis using a “reconstructed
22 vegetation index” (RVI) for the period 1901–2006 from MODIS satellite-derived NDVI (Normalised
23 Vegetation Difference Index) showed that CO₂ fertilisation contributed at least 40% of the observed increase
24 in the land carbon sink (Los 2013). Without CO₂ fertilisation ESMs are unable to simulate the increasing
25 land sink and the observed atmospheric CO₂ concentration growth rate since the middle of the 20th century
26 (Shevliakova et al. 2013). There are other mechanisms that could explain enhanced land C uptake such as
27 increased regional forest and shrub cover (see Cross-Chapter Box 2: Implications of large-scale conversion
28 from non-forest to forest land, Chapter 1);(Chen et al. 2019), and, at higher latitudes, increasing temperatures
29 and longer growing seasons (Zhu et al. 2016).

30
31 In summary, there is *low confidence* about the magnitude of the CO₂ effect and other factors that may
32 explain at least a portion of the land sink (e.g., nitrogen deposition, increased growing season, reduced
33 burning, erosion and re-deposition or organic sediments, and aerosol-induced cooling). Increases in
34 atmospheric CO₂ result in increased water use efficiency and increase leaf-level photosynthesis (*high*
35 *confidence*). The extent to which CO₂ fertilisation results in plant- or ecosystem-level carbon accumulation is
36 highly variable and affected by other environmental constraints (*high confidence*). Even in ecosystems where
37 CO₂ fertilisation has been detected in recent decades, those effects are found to weaken as a result of
38 physiological acclimation, soil nutrient limitation, and other constraints on growth (Friend et al., 2014;
39 Körner, 2006; Peñuelas et al., 2017).

40 41 42 **2.4 Emissions and impacts of short-lived climate forcers (SLCF) from land**

43
44 While the rising atmospheric concentration of GHGs is the largest driver of anthropogenic changes in
45 climate, the levels of short-lived climate forcers (SLCF) can significantly modulate regional climate by
46 altering radiation exchanges and hydrological cycle and impact ecosystems (Boucher et al. 2013; Rogelj et
47 al. 2014; Kok et al. 2018) (*high confidence*). This section assesses the current state of knowledge with
48 respect to past and future emissions of the three major SLCFs and their precursors: mineral dust,
49 carbonaceous aerosols (Black Carbon and Organic Carbon), and Biogenic volatile organic compounds
50 (BVOCs). The chapter also reports on implications of changes in their emissions for climate. Aerosols
51 particles with diameters between about 0.010 µm to about 20 µm are recognised as SLCFs, a term that refers
52 to their short atmospheric lifetime (a few days). BVOCs are important precursors of ozone and organic
53 carbon (OC), both important climate forcing agents with short atmospheric lifetimes.

54
55 While the AR5 did not assess land aerosols emissions in depth, their findings stated that although progress in
56 quantifying regional emissions of anthropogenic and natural land aerosols has been made, considerable

1 uncertainty still remains about their historical trends, their inter-annual and decadal variability and about any
2 changes in the future (Calvo et al. 2013; Klimont et al. 2017). Some new and improved understanding of
3 processes controlling emissions and atmospheric processing has been developed since AR5, for example, a
4 better understanding of the climatic role of Black Carbon (BC) as well as the understanding of the role of
5 BVOCs in formation of secondary organic aerosols (SOA).

6
7 Depending on the chemical composition and size, aerosols can absorb or scatter sunlight and thus directly
8 affect the amount of absorbed and scattered radiation (Fuzzi et al. 2015a; Nousiainen 2011; de Sá et al. 2019)
9 Aerosols affect clouds formation and development, and thus can also influence precipitation patterns and
10 amounts (Sun et al. 2015). In addition, deposition of aerosols—especially black carbon—on snow and ice
11 surfaces can reduce albedo and increase warming as a self-reinforcing feedback. Aerosols deposition also
12 change biogeochemical cycling in critical terrestrial ecosystems with deposition of nutrients such as nitrogen
13 and phosphorus (Andreae et al. 2002). Primary land aerosols are emitted directly into the atmosphere due to
14 natural or anthropogenic processes and include mineral aerosols (or dust), volcanic dust, soot from
15 combustion, organic aerosols from industry, vehicles or biomass burning, bioaerosols from forested regions,
16 and others. Secondary organic aerosols (SOA) are particulates that are formed in the atmosphere by gas-to-
17 particles conversion processes from gaseous precursors, such as BVOCs, and account for a large fraction of
18 fine mode (particles less than 2.5µm) aerosol mass (Hodzic et al. 2016; Manish et al. 2017). Land use change
19 can affect the climate through changed emissions of SLCFs such as aerosols, ozone precursors and methane.

20
21 Aerosols from air pollution will decline in the coming years as a means for improving urban and regional air,
22 but their removal will lead to additional warming (Boucher et al. 2013), with important regional variability,
23 and partially offsetting projected mitigation effects for two to three decades in 1.5°C consistent pathways
24 (IPCC 2018) (*high confidence*). It is important to emphasise that changes in emissions can either be due to
25 external forcing or through a feedback in the climate system (Box 2.1:). For instance, enhanced dust
26 emissions due to reduced vegetation could be a forcing if overgrazing is the cause of larger dust emission, or
27 a feedback if dryer climate is the cause. This distinction is important in terms of mitigation measures to be
28 implemented.

29 30 **2.4.1 Mineral dust**

31
32 One of the most abundant atmospheric aerosols emitted into the atmosphere is mineral dust, a “natural”
33 aerosol that is produced by wind strong enough to initiate the emissions process of sandblasting. Mineral
34 dust is preferentially emitted from dry and unvegetated soils in topographic depressions where deep layer of
35 alluvium have been accumulated (Prospero et al. 2002). Dust is also emitted from disturbed soils by human
36 activities with a 25% contribution to global emission, based on satellite-based estimate (Ginoux et al. 2012).
37 Dust is then transported over long distances across continents and oceans. Dust cycle, which consists of
38 mineral dust emission, transport, deposition and stabilisation, have multiple interactions with many climate
39 processes and biogeochemical cycles.

40 41 **2.4.1.1 Mineral dust as a short-lived climate forcer from land**

42 Depending on the dust mineralogy, mixing state, and size, dust particles can absorb or scatter shortwave and
43 long-wave radiation. Dust particles serve as cloud condensation nuclei and ice nuclei. They can influence the
44 microphysical properties of clouds, their lifetime and precipitation rate (Kok et al. 2018). New and improved
45 understanding of processes controlling emissions and transport of dust, its regional patterns and variability as
46 well as its chemical composition has been developed since AR5.

47
48 While satellites remain the primary source of information to locate dust sources and atmospheric burden, in-
49 situ data remains critical to constrain optical and mineralogical properties of the dust (Di Biagio et al. 2017;
50 Rocha-Lima et al. 2018). Dust particles are composed of minerals, including iron oxides which strongly
51 absorb shortwave radiation and provide nutrient for marine ecosystems. Other mineral such as feldspar is an
52 efficient ice nuclei (Harrison et al. 2016). Dust mineralogy depending on the native soils, global databases
53 were developed to characterise mineralogical composition of soils for use in the weather and climate models
54 (Journet et al. 2014; Perlwitz et al. 2015). New field campaigns as well as new analysis from prior campaign
55 have produced insights into role of dust in western Africa in climate system, for example, for dust
56 (Veselovskii et al. 2016), long-ranged transport of dust across the Atlantic (Groß et al. 2015), and the

1 characterisation of aerosol particles and their ability to act as ice and cloud condensation nuclei (Price et al.
2 2018). Size distribution at emission is another key parameter controlling dust interactions with radiation.
3 Most models use now the parametrisation of Kok (2011) based on the theory of brittle material. It was shown
4 that most models underestimate the size of global dust cycle (Kok 2011) has been underestimated.
5 Characterisation of spatial and temporal distribution of dust emissions is essential for weather prediction and
6 climate projections (*high confidence*). Although there is a growing confidence in characterising the
7 seasonality and peak of dust emissions (i.e., spring-summer, (Wang et al. 2015)) and how the meteorological
8 and soil conditions control dust sources, an understanding of long-term future dust dynamics, inter-annual
9 dust variability and how they will affect future climate still requires substantial work. Dust is also important
10 at high latitude, where it has impacts on snow covered surface albedo and weather (Bullard et al. 2016).

11 **2.4.1.2 Effects of past climate change on dust emissions, and feedbacks**

12 Limited number of model-based studies found that dust emissions increased significantly since late 19th
13 century: by 25% from preindustrial to present day (e.g., from 729 Tg yr⁻¹ to 912 Tg yr⁻¹) with ~50% increase
14 driven by climate change and ~40% by land use cover change such as conversion of natural land to
15 agriculture (Stanelle et al. 2014) (*low confidence*). These changes resulted in a clear sky radiative forcing at
16 the top of the atmosphere of -0.14 W m⁻² (Stanelle et al. 2014). The authors found that, in North Africa most
17 dust is of natural origin with a recent 15% increase in dust emissions attributed to climate change; in North
18 America two thirds of dust emissions take place on agricultural lands and both climate change and land use
19 change jointly drive the increase; between pre-industrial and present-day the overall effect of changes in dust
20 is -0.14 W m⁻² cooling of clear sky net radiative forcing on top of the atmosphere, with -0.05 W m⁻² from land
21 use and -0.083 W m⁻² from changes in climate.

22
23
24 The comparison of observations for vertically integrated mass of atmospheric dust mass per unit area (i.e.
25 Dust Mass Path or DMP) obtained from the remotely sensed data and the DMP from CMIP5 models reveal
26 that model-simulate range of DMP was much lower than the estimates from (Evan et al. 2014). ESM
27 typically do not reproduce inter-annual and longer time scales variability seen in observations (Evan et al.
28 2016). Analyses of the CMIP5 models (Evan 2018; Evan et al. 2014)) reveal that all climate models
29 systematically under-estimate dust emissions, amount of dust in the atmosphere and its inter-annual
30 variability (*medium confidence*).

31
32 One commonly suggested reason for the lack of dust variability in climate models is the models' inability to
33 simulate the effects of land surface changes on dust emission (Stanelle et al. 2014). Models which account
34 for changes in land surface show more agreement with the satellite observations both in terms of Aerosol
35 Optical Depth and DMP (Kok et al. 2014). New prognostic dust emissions models now able to account for
36 both changes in surface winds and vegetation characteristics (e.g., leaf area index and stem area index) and
37 soil water, ice, and snow cover (Evans et al. 2016). As a result, new modelling studies (e.g. Evans et al.
38 2016) indicate that in regions where soil and vegetation respond strongly to ENSO events, such as in
39 Australia, inclusion of dynamic vegetation characteristics into dust emission parameterisations improves
40 comparisons between the modelled and observed relationship long-term climate variability (e.g., ENSO) and
41 dust levels (Evans et al. 2016). Thus, there has been progress in incorporating effects of vegetation, soil
42 moisture, surface wind and vegetation on dust emission source functions but the number of studies
43 demonstrating such improvement remains small (*limited evidence, medium agreement*).

44 **2.4.1.3 Future changes of dust emissions**

45
46
47 There is no agreement about direction of future changes in dust emissions. Atmospheric dust loading is
48 projected to increase over the southern edge of the Sahara in association with surface wind and precipitation
49 changes (Pu and Ginoux, 2018), while Evan et al. (2016) project a decline in African dust emissions. Dust
50 Optical Depth (DOD) is also projected to increase over the central Arabian Peninsula in all seasons and to
51 decrease over northern China from MAM to SON (Pu and Ginoux 2018). Climate models project rising
52 drought risks over the southwestern and central US. in the twenty-first century. The projected drier regions
53 largely overlay the major dust sources in the US. However, whether dust activity in the US will increase in
54 the future is not clear, due to the large uncertainty in dust modelling (Pu and Ginoux 2017). Future trends of
55 dust emissions will depend on changes in precipitation patterns and atmospheric circulation (*limited*
56 *evidence, high agreement*). However, implication of changes in human activities, including mitigation (e.g.

1 bioenergy production) and adaption (e.g. irrigation) are not characterised in the current literature.

2 3 4 **2.4.2 Carbonaceous Aerosols**

5
6 Carbonaceous aerosols are one of the most abundant components of aerosol particles in continental areas of
7 the atmosphere and a key land-atmosphere component (Contini et al. 2018). They can make up to 60-80% of
8 PM_{2.5} (Particulate matter with size less than 2.5 µm) in urban and remote atmosphere (Tsigaridis et al.
9 2014a; Kulmala et al. 2011). It comprises an organic fraction (Organic Carbon - OC) and a refractory light
10 absorbing component, generally referred as Elemental Carbon (EC), from which Black Carbon (BC) is the
11 optically active absorption component of EC (Gilardoni et al. 2011; Bond et al. 2013).

12 13 **2.4.2.1 Carbonaceous aerosol precursors of short-lived climate forcers from land**

14 OC is a major component of aerosol mass concentration, and it originates from different anthropogenic
15 (combustion processes) and natural (from natural biogenic emissions) sources (Robinson et al. 2007). A
16 large fraction of OC in the atmosphere has a secondary origin, as it can be formed in the atmosphere through
17 condensation to the aerosol phase of low vapour pressure gaseous compounds emitted as primary pollutants
18 or formed in the atmosphere. This component is called Secondary Organic Aerosol (SOA) (Hodzic et al.
19 2016). A third component of the optically active aerosols is the so-called brown carbon (BrC), an organic
20 material that shows enhanced solar radiation absorption at short wavelengths (Wang et al. 2016b; Laskin et
21 al. 2015; Liu et al. 2016a; Bond et al. 2013; Saturno et al. 2018).

22
23 OC and EC have distinctly different optical properties, with OC being important for the scattering properties
24 of aerosols and EC is central for the absorption component (Rizzo et al. 2013; Tsigaridis et al. 2014a; Fuzzi
25 et al. 2015a). While organic carbon is reflective and scatter solar radiation, it has a cooling effect on climate.
26 On the other side, BC and BrC absorbs solar radiation and they have a warming effect in the climate system.
27 (Bond et al. 2013).

28
29 Organic carbon is also characterised by a high solubility with a high fraction of water-soluble organic
30 compounds (WSOC) and it is one of the main drivers of the oxidative potential of atmospheric particles. This
31 makes particles loaded with oxidised OC an efficient CCN in most of the conditions (Pöhlker et al. 2016;
32 Thalman et al. 2017; Schmale et al. 2018).

33
34 Biomass burning is a major global source of carbonaceous aerosols (Bowman et al. 2011b; Harrison et al.
35 2010; Reddington et al. 2016; Artaxo et al. 2013). As knowledge of past fire dynamics improved through
36 new satellite observations, new fire proxies' datasets (Marlon et al. 2013; van Marle et al. 2017), and
37 process-based models (Hantson et al. 2016), a new historic biomass burning emissions dataset starting in
38 1750 has been developed (Van Marle et al. 2017a) (see Cross-Chapter Box 3: Fire and Climate Change, in
39 this chapter). Revised versions of OC biomass burning emissions (Van Marle et al. 2017a) show in general
40 reduced trends compared to the emissions derived by (Lamarque et al. 2010) for CMIP5. CMIP6 global
41 emissions pathways (Gidden et al. 2018; Hoesly et al. 2018) estimate global BC emissions in 2015 at 9.8 Mt
42 BC yr⁻¹, while global OC emissions are 35 Mt OC yr⁻¹.

43
44 Land use change is critically important for carbonaceous aerosols, since biomass burning emissions consist
45 mostly of organic aerosol, and the undisturbed forest is also a large source of organic aerosols (Artaxo et al.
46 2013). Additionally, urban aerosols are also mostly carbonaceous, because of the source composition (traffic,
47 combustion, industry, etc.) (Fuzzi et al. 2015b). Burning of fossil fuel, biomass burning emissions and SOA
48 from natural BVOC emissions are the main global sources of carbonaceous aerosols. Any change in each of
49 these components influence directly the radiative forcing (Contini et al. 2018; Boucher et al. 2013; Bond et
50 al. 2013).

51
52 One important component of carbonaceous aerosols is the primary biological aerosol particles (PBAP), also
53 called bioaerosols, that correspond to a significant fraction of aerosols in forested areas (Fröhlich-Nowoisky
54 et al. 2016; Pöschl and Shiraiwa 2015). They are emitted directly by the vegetation as part of the biological
55 processes (Huffman et al. 2012). Airborne bacteria, fungal spores, pollen, archaea, algae, and other
56 bioparticles are essential for the reproduction and spread of organisms across various terrestrial ecosystems.

1 They can serve as nuclei for cloud droplets, ice crystals, and precipitation, thus influencing the hydrological
2 cycle and climate (Whitehead et al. 2016; Scott et al. 2015; Pöschl et al. 2010).

3 4 **2.4.2.2 Effects of past climate change on carbonaceous aerosols emissions, and feedbacks**

5 Annual global emission estimates of BC range from 7.2-7.5 Tg yr⁻¹ using bottom-up inventories (Bond et al.
6 2013; Klimont et al. 2017) up to 17.8 ± 5.6 Tg yr⁻¹ using a fully coupled climate-aerosol-urban model
7 constrained by aerosol measurements (Cohen and Wang 2014), with considerably higher BC emissions for
8 Eastern Europe, Southern East Asia, and Southeast Asia mostly due to higher anthropogenic BC emissions
9 estimates. A significant source of BC, the net trend in global burned area from 2000 to 2012 was a modest
10 decrease of 4.3 Mha yr⁻¹ (-1.2% yr⁻¹).

11
12 Carbonaceous aerosols are important in urban areas as well as pristine continental regions, since they can be
13 responsible for 50-85% of PM_{2.5} (Contini et al. 2018; Klimont et al. 2017). In boreal and tropical forests,
14 carbonaceous aerosols originate from BVOC oxidation (Section 2.4.3). The largest global source of BC
15 aerosols is open burning of forests, savannah and agricultural lands with emissions of about 2,700 Gg yr⁻¹
16 in the year 2000 (Bond et al. 2013).

17
18 ESMS most likely underestimate globally averaged EC emissions (Bond et al. 2013; Cohen and Wang 2014),
19 although recent emission inventories have included an upwards adjustment in these numbers (Hoesly et al.
20 2018). Vertical EC profiles have also been shown to be poorly constrained (Samset et al. 2014a), with a
21 general tendency of too much EC at high altitudes. Models differ strongly in the magnitude and importance
22 of the coating-enhancement of ambient EC absorption (Boucher et al. 2016) (Gustafsson and Ramanathan
23 2016), in their estimated lifetime of these particles, as well as in dry and wet removal efficiency (Mahmood
24 et al. 2016) (*limited evidence, medium agreement*).

25
26 The equilibrium in emissions and concentrations between the scattering properties of organic aerosol versus
27 the absorption component of BC is a key ingredient in the future climatic projections of aerosol effects
28 (*limited evidence, high agreement*). The uncertainties in net climate forcing from BC rich sources are
29 substantial, largely due to lack of knowledge about cloud interactions with both black carbon and co-emitted
30 organic carbon. A strong positive forcing of about 1.1 W m⁻² was calculated by (Bond et al. 2013), but this
31 forcing is balanced by a negative forcing of -1.45 W m⁻², and shows clearly a need to work on the co-
32 emission issue for carbonaceous aerosols. The forcing will also depend on the aerosol-cloud interactions,
33 where carbonaceous aerosol can be coated and change their CCN capability. It is difficult to estimate the
34 changes in any of these components in a future climate, but this will influence strongly the radiative forcing
35 (Contini et al. 2018; Boucher et al. 2013; Bond et al. 2013) (*high confidence*).

36
37 De Coninck et al. (2018) reported studies estimating a lower global temperature effect from BC mitigation
38 (e.g., Samset et al. 2014b; Boucher et al. 2016), although commonly used models do not capture properly
39 observed effects of BC and co-emissions on climate (e.g., (Bond et al. 2013). Regionally, the warming
40 effects can be substantially larger, for example, in the Arctic (Sand et al. 2015) and high mountain regions
41 near industrialised areas or areas with heavy biomass burning impacts (Ming et al. 2013) (*high confidence*).

42 43 **2.4.2.3 Future changes of carbonaceous aerosol emissions**

44 Due to the short atmospheric lifetime of carbonaceous aerosols in the atmosphere, of the order of a few days,
45 most studies dealing with the future concentration levels have a regional character (Cholakian et al. 2018;
46 Fiore et al. 2012). The studies agree that the uncertainties in changes in emissions of aerosols and their
47 precursors are generally higher than those connected to climate change itself. Confidence in future changes
48 in carbonaceous aerosol concentration projections is limited by the reliability of natural and anthropogenic
49 emissions (including wildfires, largely caused by human activity) of primary aerosol as well as that of the
50 precursors. The Aerosol Chemistry Model Intercomparison Project (AerChemMIP) is endorsed by the
51 Coupled-Model Intercomparison Project 6 (CMIP6) and is designed to quantify the climate impacts of
52 aerosols and chemically- reactive gases (Lamarque et al. 2013). These simulations calculated future
53 responses to SLCF emissions for the RCP scenarios in terms of concentration changes and radiative forcing.
54 Carbonaceous aerosol emissions are expected to increase in the near future due to possible increases in open
55 biomass burning (forest, savannah, and agricultural fires) emissions, and increase in SOA from oxidation of
56 BVOCs (Tsigaridis et al. 2014b; Van Marle et al. 2017b; Giglio et al. 2013) (*medium confidence*).

1
2 More robust knowledge has been produced since the conclusions reported in AR5 (Boucher et al. 2013) and
3 all lines of evidence now agree on a small effect on carbonaceous aerosol global burden due to climate
4 change (*medium confidence*). The regional effects, however, are predicted to be much higher (Westervelt et
5 al. 2015). With respect to possible changes in the chemical composition of PM as a result of future climate
6 change only a few sparse data are available in the literature and the results are, as yet, inconclusive. The co-
7 benefits of reducing aerosol emissions due to air quality issues will play an important role in future
8 carbonaceous aerosol emissions (Gonçalves et al. 2018; Shindell et al. 2017) (*high confidence*).
9

10 **2.4.3 Biogenic Volatile Organic Compounds (BVOCs)**

11
12 Biogenic volatile organic compounds (BVOCs) are emitted in large amounts by forests (Guenther et al.
13 2012). They include isoprene, terpenes, alkanes, alkenes, alcohols, esters, carbonyls and acids (Peñuelas and
14 Staudt 2010; Guenther et al. 1995, 2012). Their emissions represent a carbon loss to the ecosystem, which
15 can represent up to 10% of the carbon fixed by photosynthesis under stressful conditions (Bracho-Nunez et
16 al. 2011). The global average emission for vegetated surfaces is $0.7 \text{ g C m}^{-2} \text{ yr}^{-1}$ but can exceed 100 g C m^{-2}
17 yr^{-1} in some tropical ecosystems (Peñuelas and Llusà 2003).
18

19 **2.4.3.1 BVOC precursors of short-lived climate forcers from Land**

20 BVOCs are rapidly oxidised in the atmosphere to form less volatile compounds that can condense and form
21 secondary organic aerosol (SOA). In boreal and tropical forests, carbonaceous aerosols originate from
22 BVOC oxidation, of which isoprene and terpenes are the most important precursors (Claeys et al. 2004; Hu
23 et al. 2015; De Sá et al. 2017; de Sá et al. 2018; Liu et al. 2016b, see following sub-section). BVOCs are the
24 most important precursors of SOA. This transformation process of BVOCs affects the aerosol size
25 distribution both by contributing to new particle formation and to the growth of larger pre-existing particles.
26 SOA affect the scattering of radiation by the particles themselves (direct aerosol effect), but also change the
27 amount of cloud condensation nuclei (CCN) and the lifetime and optical properties of clouds (indirect
28 aerosol effect).
29

30 High amounts of SOA are observed over forest areas, in particular in boreal and tropical regions where they
31 have been found to mostly originate from BVOC emissions (Manish et al. 2017). In particular, isoprene
32 epoxydiol-derived SOA (IEPOX-SOA) is being identified in recent studies in North America and
33 Amazonian forest as a major component in the oxidation of isoprene (Allan et al. 2014; Schulz et al. 2018;
34 De Sá et al. 2017). In tropical regions BVOC can be convected up to the upper atmosphere, where their
35 volatility is reduced and where they become SOA. In some cases those particles are even transported back to
36 the lower atmosphere (Schulz et al. 2018; Wang et al. 2016a; Andreae et al. 2018). In the upper troposphere
37 in the Amazon, SOA are important CCN and are responsible for the vigorous hydrological cycle (Pöhlker
38 et al. 2018). This strong link between BVOC emissions by plants and hydrological cycle has been discussed in
39 a number of studies (Fuentes et al. 2000; Schmale et al. 2018; Pöhlker et al. 2018, 2016).
40

41 Changing BVOC emissions also affect the oxidant concentrations in the atmosphere. Their impact on the
42 concentration of ozone depends on the NO_x concentrations. In polluted regions, high BVOC emissions lead
43 to increased production of ozone, followed by the formation of more OH and a reduction in the methane
44 lifetime. In more pristine regions (NO_x-limited), increasing BVOC emissions instead lead to decreasing OH
45 and ozone concentrations, resulting in a longer methane lifetime. The net effect of BVOCs then can change
46 over time if NO_x emissions are changing.
47

48 BVOCs' possible climate effects have received little attention because it was thought that their short lifetime
49 would preclude them from having any significant direct influence on climate (Unger 2014a; Sporre et al.
50 2018). Higher temperatures and increased CO₂ concentrations are (separately) expected to increase the
51 emissions of BVOCs (Jardine et al. 2011, 2015; Fuentes et al. 2016). This has been proposed to initiate
52 negative climate feedback mechanisms through increased formation of SOA (Arneth et al. 2010; Kulmala
53 2004; Unger et al. 2017). More SOA can make the clouds more reflective, which can provide a cooling.
54 Furthermore, the increase in SOA formation has also been proposed to lead to increased aerosol scattering,
55 resulting in an increase in diffuse radiation. This could boost gross primary production (GPP) and further
56 increase BVOC emissions (Kulmala et al. 2014; Cirino et al. 2014; Sena et al. 2016; Schafer et al. 2002;

Ometto et al. 2005; Oliveira et al. 2007). These important feedbacks are starting to emerge (Sporre et al. 2018; Kulmala 2004; Arneth et al. 2017b). However, there is evidence that this influence might be significant at different spatial scales, from local to global, through aerosol formation and through direct and indirect greenhouse effects (*limited evidence, medium agreement*). Most tropical forest BVOC are primarily emitted from foliage of trees but soil microbes can also be a major source of some compounds including sesquiterpenes (Bourtsoukidis et al. 2018).

2.4.3.2 *Historical changes of BVOCs and contribution to climate change*

Climate warming over the past 30 years, together with the longer growing season experienced in boreal and temperate environments, have increased BVOC global emissions by since the preindustrial times (*limited evidence, medium agreement*) (Peñuelas 2009; Sanderson et al. 2003; Pacifico et al. 2012). This was opposed by lower BVOC emissions caused by the historical conversion of natural vegetation and forests to cropland (*limited evidence, medium agreement*) (Unger 2013, 2014a; Fu and Liao 2014). The consequences of historical anthropogenic land cover change were a decrease in the global formation of SOA (-13 %, Scott et al. 2017) and tropospheric burden (-13 %, Heald and Geddes 2016). This has resulted in a positive radiative forcing (and thus warming) from 1850 to 2000 of 0.017 W m^{-2} (Heald and Geddes 2016), 0.025 (Scott et al. 2017) and 0.09 W m^{-2} (Unger 2014b) through the direct aerosol effect. In present-day conditions, global SOA production from all sources spans between 13 and 121 Tg yr^{-1} (Tsigaridis et al. 2014a). The indirect aerosol effect (change in cloud condensation nuclei), resulting from land use induced changes in BVOC emissions, adds an additional positive radiative forcing of 0.008 W m^{-2} (Scott et al. 2017). More studies with different model setups are needed to fully assess this indirect aerosol effect associated with land use change from the preindustrial to present. CMIP6 global emissions pathways (Hoesly et al. 2018; Gidden et al. 2018) estimates global VOCs emissions in 2015 at $230 \text{ Mt VOC yr}^{-1}$. They also estimated that from 2000 to 2015, emissions were up from 200 to $230 \text{ Mt VOC yr}^{-1}$.

There is (*limited evidence, medium agreement*) that historical changes in BVOC emissions have also impacted tropospheric ozone. At most surface locations where land use has changed, the NO_x concentrations are sufficiently high for the decrease in BVOC emissions to lead to decreasing ozone concentrations (Scott et al. 2017). However, in more pristine regions (with low NO_x concentrations), the imposed conversion to agriculture has increased ozone through decreased BVOC emissions and their subsequent decrease in OH (Scott et al. 2017; Heald and Geddes 2016). In parallel, the enhanced soil NO_x emissions from agricultural land, can increase the ozone concentrations in NO_x limited regions (Heald and Geddes 2016).

Another impact of historical decrease in BVOC emissions is the reduction in the atmospheric lifetime of methane (*limited evidence, medium agreement*), which results in a negative radiative forcing that ranges from -0.007 W m^{-2} (Scott et al. 2017) to -0.07 W m^{-2} (Unger 2014b). However, the knowledge of to which degree BVOC emissions impact oxidant concentrations, in particular OH (and thus methane concentrations), is still limited and therefore these numbers are very uncertain (Heald and Spracklen 2015; Scott et al. 2017). The effect of land use change on BVOC emissions are highly heterogeneous (Rosenkranz et al. 2015) and though the global values of forcing described above are small, the local or regional values can be higher and even of opposite sign than the global values.

2.4.3.3 *Future changes of BVOCs*

Studies suggest that increasing temperature will change BVOC emissions through change in species composition and rate of BVOC productions. A further 2°C to 3°C rise in the mean global temperature, could increase BVOC global emissions by an additional 30–45% (Peñuelas and Llusà 2003). In two modelling studies, the impact on climate from rising BVOC emissions have been found to become even larger with decreasing anthropogenic aerosol emissions (Kulmala et al. 2013; Sporre et al. 2018). A negative feedback on temperature, arising from the BVOC-induced increase in the first indirect aerosol effect have been estimated by two studies to be in the order of $-0.01 \text{ W m}^{-2} \text{ K}$ (Scott et al. 2018b; Paasonen et al. 2013). Enhanced aerosol scattering from increasing BVOC emissions has been estimated to contribute with a global gain in BVOC emissions of 7% (Rap et al. 2018). In a warming planet, BVOC emissions are expected to increase but magnitude is unknown and will depend on future land use change, in addition to climate (*limited evidence, medium agreement*).

There is a very limited number of studies investigating the climate impacts of BVOCs using future land use

scenarios (Ashworth et al. 2012; Pacifico et al. 2012). Scott et al. (2018a) found that a future deforestation according to the land use scenario in RCP8.5 leads to a 4% decrease in BVOC emissions at the end of the century. This resulted in a direct aerosol forcing of $+0.006 \text{ W m}^{-2}$ (decreased reflection by particles in the atmosphere) and a first indirect aerosol forcing of -0.001 W m^{-2} (change in the amount of CCN). Studies not including future land use scenarios but investigating the climate feedbacks leading to increasing future BVOC emissions, have found a direct aerosol effect of -0.06 W m^{-2} (Sporre et al. 2018) and an indirect aerosol effect of -0.45 W m^{-2} (Makkonen et al. 2012; Sporre et al. 2018). The stronger aerosol effects from the feedback compared to the land use are, at least partly, explained by a much larger change in the BVOC emissions.

A positive climate feedback could happen in a future scenario with increasing BVOC emissions, where higher ozone and methane concentrations could lead to an enhanced warming which could further increase BVOC emissions (Arneeth et al. 2010). This possible feedback is mediated by NO_x levels. One recent study including dynamic vegetation, land use change, CO₂ and climate change found no increase or even a slight decrease in global BVOC emissions at the end of the century (Hantson et al. 2017). There is a lack of understanding concerning the processes governing the BVOC emissions, the oxidation processes in the atmosphere, the role of the BVOC oxidation products in new particle formation and particle growth, as well as general uncertainties in aerosol-cloud interactions. There is a need for continued research into these processes but the current knowledge indicates that changing BVOC emissions need to be taken into consideration when assessing the future climate and how land use will affect it. In summary, the magnitude and sign of net effect of BVOC emissions on the radiation budget and surface temperature is highly uncertain.

2.5 Land impacts on climate and weather through biophysical and GHGs effects

The focus of this section is summarised **Figure 2.13**. We report on what we know regarding the influence land has on climate via biophysical and biogeochemical exchanges. Biogeochemical effects herein only refer to changes in net emissions of CO₂ from land. The influence of land on atmospheric composition is discussed in Section 2.3.

All sections discuss impacts of land on global and regional climate, and climate extremes, whenever the information is available. Section 2.5.1 presents effects of historical and future land use scenarios; section 2.5.2 is devoted to impacts of specific anthropogenic land uses such as forestation, deforestation, irrigation, crop and forest management; section 2.5.3 focuses on how climate driven land changes feedback on climate and section 2.5.4 puts forward that land use changes in one region can affect another region.

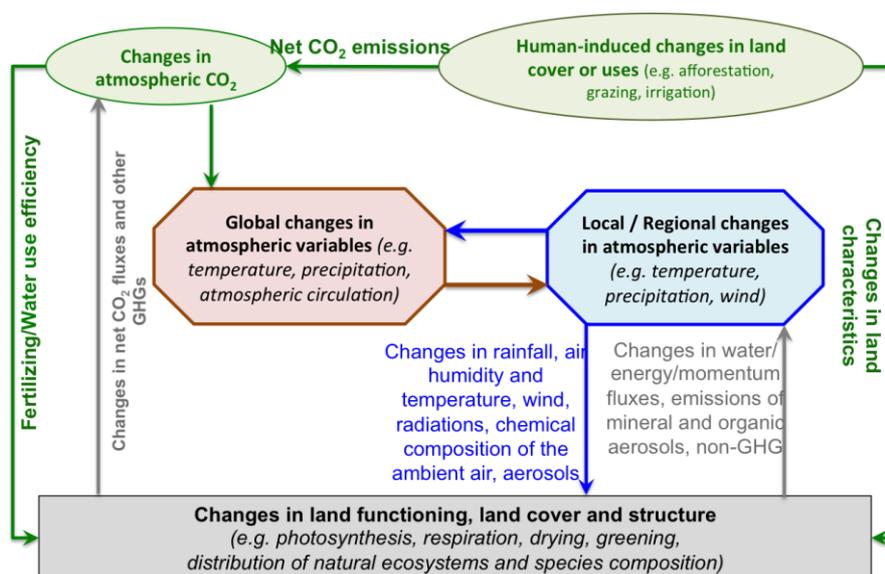


Figure 2.13 Global, local and regional climate changes are the focus of this section. They are examined

through changes in climate states (e.g., changes in air temperature and humidity, rainfall, radiation) as well as through changes in atmospheric dynamics (e.g., circulation patterns). Changes in land that influence climate are either climate- or Human- driven. Green arrows and boxes refer to what we consider herein as imposed changes (forcings). Grey box and arrows refer to responses of land to forcings (green and blue boxes) and feedbacks on those initial forcings. Red and blue boxes and arrows refer respectively to global and local/regional climate changes and their subsequent changes on land

2.5.1 Impacts of historical and future anthropogenic land cover changes

The studies reported below focus essentially on modelling experiments, as there is no direct observation of how historical land use changes have affected the atmospheric dynamics and physics at the global and regional scales. Moreover, the climate modelling experiments only assess the impacts of anthropogenic land cover changes (e.g. deforestation, urbanisation) and neglect the effects of changes in land management (e.g. irrigation, use of fertilisers, choice of species varieties among managed forests or crops). Because of this restricted accounting for land use changes we will use the term land cover changes in the following sub-sections (2.5.1.1 and 2.5.1.2).

Each section starts by describing changes at the global scale, and at the regional scale and ends with what we know about the impacts of those scenarios on extreme weather events, whenever the information is available.

2.5.1.1 Impacts of global historical land cover changes on climate

2.5.1.1.1 At the global level

The contribution of anthropogenic land cover changes to the net global warming throughout the 20th century has been derived from few model-based estimates that account simultaneously for biogeochemical and biophysical effects of land on climate (Table 2.4). The simulated net change in mean global annual surface air temperature, averaged over all the simulations, is a small warming of $0.078 \pm 0.093^\circ\text{C}$, ranging from small cooling simulated by two models (-0.05°C and -0.02°C respectively in (Brovkin et al. 2004) and (Simmons and Matthews 2016), to larger warming simulated by three models ($>+0.14^\circ\text{C}$ (Shevliakova et al. 2013; Pongratz et al. 2010; Matthews et al. 2004). When starting from the Holocene period (He et al. 2014) has estimated an even larger net warming effect of anthropogenic land cover changes ($+0.72^\circ\text{C}$).

Table 2.4 Change in mean global annual surface air temperature resulting from anthropogenic land cover change over the historical period. This historical period varies from one simulation to another (middle column).

| Reference of the study | Time period | Mean global annual change in surface air temperature ($^\circ\text{C}$) |
|---|-------------|---|
| (Simmons and Matthews 2016) | 1750-2000 | -0.02 |
| (Shevliakova et al. 2013) | 1861-2005 | +0.17 |
| (Pongratz et al. 2010) | 1900-2000 | +0.14 |
| (Matthews et al. 2004) | 1700-2000 | +0.15 |
| (Brovkin et al. 2004) | 1850-2000 | -0.05 |
| Mean \pm standard deviation | | 0.078 ± 0.093 |

This net small warming signal results from the competing effects of biophysical cooling (*medium confidence*) and biogeochemical warming (*very high confidence*; Figure 2.14²). The global biophysical cooling alone has been estimated by a larger range of climate models and is $-0.10 \pm 0.14^\circ\text{C}$; it ranges from -0.57°C to $+0.06^\circ\text{C}$ (e.g. (Zhang et al. 2013a; Hua and Chen 2013; Jones et al. 2013b; Simmons and Matthews 2016), Table A2.1). This cooling is essentially dominated by increases in surface albedo: historical land cover changes have generally led to a dominant brightening of land as discussed in AR5 (Myhre et al. 2013). Reduced incoming long-wave radiation at the surface from reduced evapotranspiration and thus less water vapour in the atmosphere has also been reported as a potential contributor to this cooling (Claussen et al. 2001a). The cooling is however dampened by decreases in turbulent fluxes leading to decreased loss of

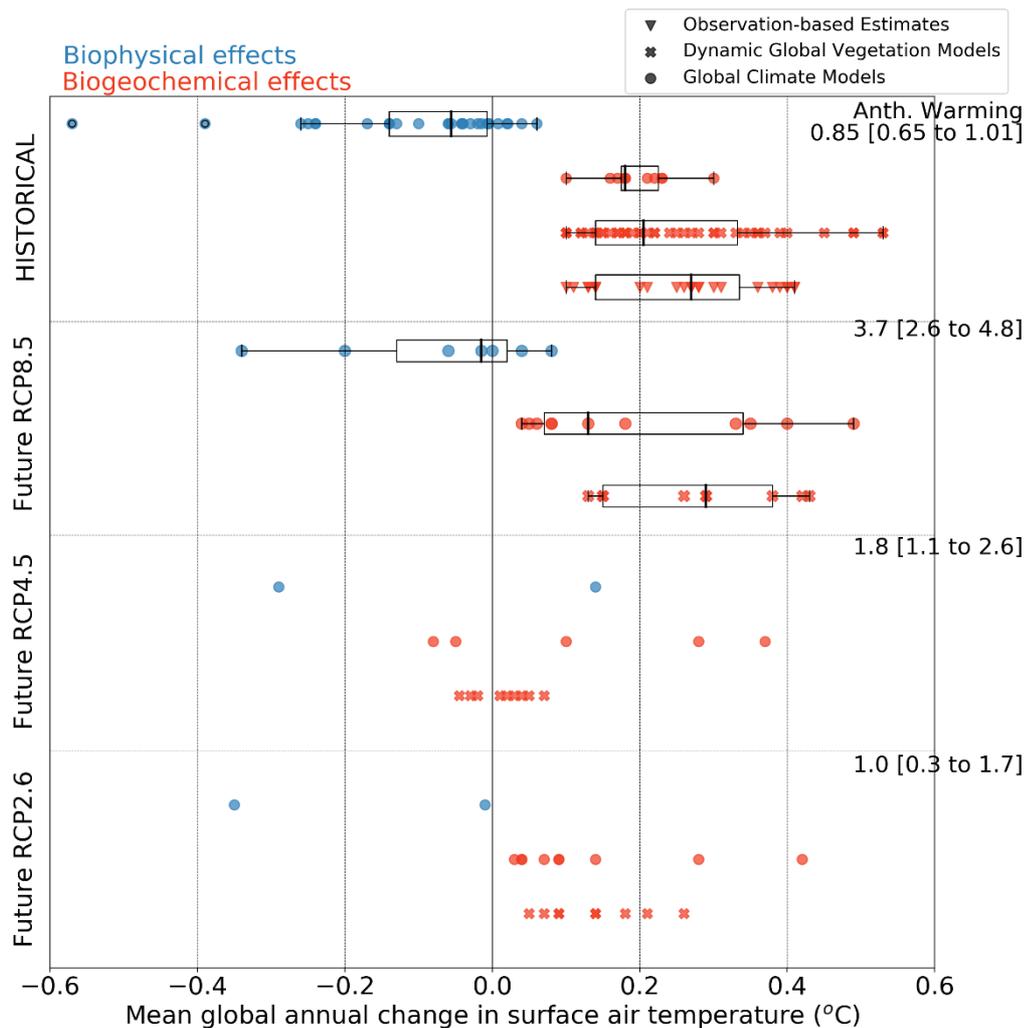
² FOOTNOTE: The detailed list of all values used to construct this figure is provided in Table A2.1 in the Appendix at the end of the Chapter

1 heat and water vapour from the land through convective processes. Those non-radiative processes are indeed
2 well-known to often oppose the albedo-induced surface temperature changes (e.g., (Davin and de Noblet-
3 Ducoudre 2010; Boisier et al. 2012)).

4
5 Historical land cover changes have contributed to the increase in atmospheric CO₂ content (Section 2.3) and
6 thus to global warming (biogeochemical effect, *very high confidence*). The global mean biogeochemical
7 warming has been calculated from observation-based estimates (+0.25±0.10°C; e.g. (Li et al. 2017a;
8 Avitabile et al. 2016; Carvalhais et al. 2014; Le Quéré et al. 2015)), or estimated from dynamic global
9 vegetation models (+0.24±0.12°C; e.g. (Peng et al. 2017; Arneth et al. 2017a; Pugh et al. 2015; Hansis et al.
10 2015)) and global climate models (+0.20±0.05°C; (Pongratz et al. 2010; Brovkin et al. 2004; Matthews et al.
11 2004; Simmons and Matthews 2016)).

12
13 The magnitude of these simulated biogeochemical effects may however be underestimated as they do not
14 account for a number of processes such as land management, nitrogen/phosphorus cycles, changes in the
15 emissions of CH₄, N₂O and non-GHG emissions from land (Ward et al. 2014; Arneth et al. 2017b; Cleveland
16 et al. 2015; Pongratz et al. 2018). Two studies have accounted for those compounds and found a global net
17 positive radiative forcing in response to historical anthropogenic land cover changes, indicating a net surface
18 warming (Mahowald et al. 2017; Ward et al. 2014). However, first the estimated biophysical radiative
19 forcing in those studies only accounts for changes in albedo and not for changes in turbulent fluxes.
20 Secondly, the combined estimates also depend on other several key modelling estimates such as climate
21 sensitivity, CO₂ fertilisation caused by land use emissions, possible synergistic effects, validity of radiative
22 forcing concept for land forcing. The comparison with the other above-mentioned modelling studies is thus
23 difficult.

24
25 In addition, most of those estimates do not account for the evolution of natural vegetation in unmanaged
26 areas, while observations and numerical studies have reported a greening of the land in boreal regions
27 resulting from both extended growing season and poleward migration of tree lines (Lloyd et al. 2003; Lucht
28 et al. 1995), Section 2.2). This greening enhances global warming via a reduction of surface albedo (winter
29 darkening of the land through the snow-albedo feedbacks, e.g. (Forzieri et al. 2017)). At the same time
30 cooling occurs due to increased evapotranspiration during the growing season, along with enhanced
31 photosynthesis, i.e. increased CO₂ sink (Qian et al. 2010). When feedbacks from the poleward migration of
32 treeline is accounted for together with the biophysical effects of historical anthropogenic land cover change,
33 the biophysical annual cooling (about -0.20°C to -0.22°C on land, -0.06°C globally) is significantly
34 dampened by the warming (about +0.13°C) resulting from the movements of natural vegetation (Strengers et
35 al. 2010). Accounting simultaneously for both anthropogenic and natural land cover changes reduces the
36 cooling impacts of historical land cover change in this specific study.
37



1
2 **Figure 2.14** Changes in mean global annual surface air temperature (°C) in response to historical and
3 **future anthropogenic land cover changes** as estimated from a range of studies (see Table A2.1 in the
4 **Appendix for detailed information**). Temperature changes resulting from biophysical processes (e.g.
5 **changes in physical land surface characteristics such as albedo, evapotranspiration, and roughness**
6 **length) are illustrated using blue symbols**; temperature changes resulting from biogeochemical processes
7 **(e.g. changes in atmospheric CO₂ composition) use red symbols**. Future changes are shown for three
8 **distinct scenarios: RCP8.5, RCP4.5 & RCP2.6**. The markers ‘filled circle’, ‘filled cross’, and ‘filled
9 **triangle down**’ represent estimates from respectively global climate models, dynamic global vegetation
10 **models (DGVMs), and observations**. When the number of estimate is sufficiently large, box plots are
11 **overlaid; they show the ensemble minimum, first quartile (25th percentile), median, third quartile (75th**
12 **percentile), and the ensemble maximum**. Scatter points beyond the box plot are the outliers. Details
13 **about how temperature change is estimated from DGVMs and observations is provided in Appendix**.
14 **Numbers on the right hand-side give the mean and the range of simulated mean global annual warming**
15 **from various climate models**.

16
17 **2.5.1.1.2 At the regional level**

18 The global and annual estimates reported above mask out very contrasted regional and seasonal differences.
19 Biogeochemical effects of anthropogenic land cover change on temperature follow the spatial patterns of
20 GHG-driven climate change with stronger warming over land than ocean, and stronger warming in northern
21 high latitudes than in the tropics and equatorial regions (Arctic amplification). Biophysical effects on the
22 contrary are much stronger where land cover has been modified than in their surroundings (see 2.5.4 for a
23 discussion on non-local effects). Very contrasted regional temperature changes can thus result depending on
24 whether biophysical processes dampen or exacerbate biogeochemical impacts.
25

Figure 2.15 compares, for seven climate models, the biophysical effects of historical anthropogenic land cover change in North America and Eurasia (essentially cooling) to the regional warming resulting from the increased atmospheric CO₂ content since pre-industrial times ((De Noblet-Ducoudré et al. 2012); comparing 1973–2002 to 1871–1900). It shows a dominant biophysical cooling effect of changes in land cover, at all seasons, as large as the regional footprint of anthropogenic global warming. Averaged over all agricultural areas of the world (Pongratz et al. 2010) reported a 20th century biophysical cooling of -0.10°C, and (Strengers et al. 2010) reported a land induced cooling as large as -1.5°C in western Russia and eastern China between 1871 and 2007. There is thus *medium confidence* that anthropogenic land cover change has dampened warming in many regions of the world over the historical period.

Very few studies have explored the effects of historical land cover changes on seasonal climate. There are however evidences that the seasonal magnitude and sign of those effects at the regional level are strongly related to soil-moisture/evapotranspiration and snow regimes, particularly in temperate and boreal latitudes (Teuling et al. 2010; Pitman and de Noblet-Ducoudré 2012; Alkama and Cescatti 2016). Quesada et al. (2017a) showed that atmospheric circulation changes can be significantly strengthened in winter for tropical and temperate regions. However, the lack of studies underlines the need for a more systematic assessment of seasonal, regional and other than mean temperature metrics in the future.

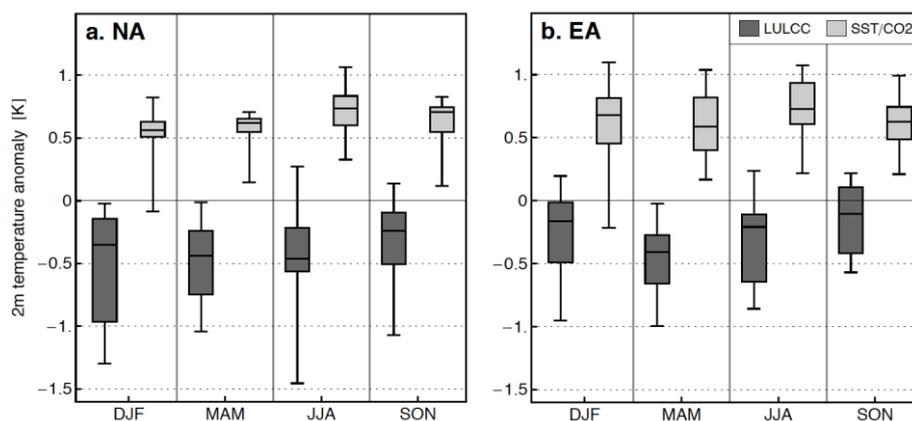


Figure 2.15 Simulated changes in mean surface air temperature (°C) between the pre-industrial period (1870–1900) and present-day (1972–2002) for all seasons and for a) North America and b) Eurasia (De Noblet-Ducoudré et al. 2012). Light grey boxes are the changes simulated in response to increased atmospheric GHG content between both time periods and subsequent changes in sea-surface temperature and sea-ice extent (SST/CO₂); the CO₂ changes accounted for include emissions from all sources including land use. Dark grey boxes are the changes simulated in response to the biophysical effects of historical land cover changes. The box-and-whisker plots have been drawn using results from seven climate models and ensembles of ten simulations per model and time period. The bottom and top of the each grey box are the 25th and 75th percentiles, and the horizontal line within each box is the 50th percentile (the median). The whiskers (straight lines) indicate the ensemble maximum and minimum values. Seasons are respectively December-January-February (DJF), March-April-May (MAM), June-July-August (JJA) and September-October-November (SON). North America and Eurasia are extended regions where land use changes are the largest between the two time periods considered (their contours can be found in Figure 1 of (De Noblet-Ducoudré et al. 2012)).

2.5.1.1.3 Effects on extremes

The effect of historical deforestation on extreme temperature trends is intertwined with the effect of other climate forcings thus making it difficult to quantify based on observations. Based on results from four climate models, the impact of historical anthropogenic land cover change on temperature and precipitation extremes was found to be locally as important as changes arising from increases in atmospheric CO₂ and sea-surface temperatures, but with a lack of model agreement on the sign of changes (Pitman et al. 2012). In some regions the impact of land cover change masks or amplifies the effect of increased CO₂ on extremes (Avila et al. 2012; Christidis et al. 2013). Using an observational constraint for the local biophysical effect of land cover change applied to a set of CMIP5 climate models, (Lejeune et al. 2018) found that historical deforestation increased extreme hot temperatures in northern mid-latitudes. The results also indicate a stronger impact on the warmest temperatures compared to mean temperatures. Findell et al. (2017) reached similar conclusions, although using only a single climate model. Importantly, the climate models involved in

1 these three studies did not consider the effect of management changes which have been shown to be
2 important, as discussed Section 2.5.2.

3
4 Based on the studies discussed above there is yet *limited evidence* but *high agreement* that land cover change
5 affects local temperature extremes more than mean values. Observational studies assessing the role of land
6 cover on temperature extremes are still very limited (Zaitchik et al. 2006; Renaud and Rebetez 2008), but
7 suggest that trees dampen seasonal and diurnal temperature variations at all latitudes and even more so in
8 temperate regions compared to short vegetation (Chen et al. 2018; Duveiller et al. 2018b; Li et al. 2015a; Lee
9 et al. 2011). Furthermore, trees also locally dampen the amplitude of heat extremes (Renaud and Rebetez
10 2008; Zaitchik et al. 2006) although this result depends on the forest type, coniferous trees providing less
11 cooling effect than broadleaf trees (Renaud et al. 2011; Renaud and Rebetez 2008).

12 13 **2.5.1.2 Impacts of future global land cover changes on climate**

14 **2.5.1.2.1 At the global level**

15 The most extreme emissions scenario (RCP8.5) that has been developed for the last coordinated modelling
16 intercomparison of climate models (CMIP5) is the one that has received the most attention in the literature
17 with respect to how projected future anthropogenic land cover changes (Hurtt et al. 2011) will affect the very
18 large simulated global warming.

19
20 Seven model-based studies have examined both the biophysical and biogeochemical effects of anthropogenic
21 changes in land-cover, as projected in RCP8.5, on future climate change (Table 2.5 ; (Simmons and
22 Matthews 2016; Davies-Barnard et al. 2014; Boysen et al. 2014)). They all agree on a biogeochemical
23 warming, ranging from +0.04°C to +0.35°C, in response to land cover change. Two models predict an
24 additional biophysical warming, while the others agree on a biophysical cooling that dampens (or overrules)
25 the biogeochemical warming. Using a wider range of global climate models, the biogeochemical warming
26 (*high confidence*) is +0.20±0.15°C whereas it is +0.28±0.11°C when estimated from dynamic global
27 vegetation models (Pugh et al. 2015; Stocker et al. 2014). This biogeochemical warming is compensated for
28 by a biophysical cooling (*medium confidence*) of -0.10±0.14°C (Quesada et al. 2017a; Davies-Barnard et al.
29 2015; Boysen et al. 2014). The estimates of temperature changes resulting from anthropogenic land cover
30 changes alone remain very small compared to the projected mean warming of +3.7°C by the end of the 21st
31 century (ranging from 2.6 to 4.8°C depending on the model and compared to 1986-2005; Figure 2.14).

32
33 **Table 2.5 Change in mean global annual surface air temperature resulting from**
34 **anthropogenic land cover changes projected for the future, according to three different**
35 **scenarios: RCP8.5, RCP4.5 and RCP2.6. Temperature change resulting from biophysical**
36 **and biogeochemical effects of land cover change are examined.**

| Reference of the study | Time period | Mean global annual change in surface air temperature (°C) | | |
|------------------------------|-------------|---|----------------|---|
| | | Biophysical / Biogeochemical | | |
| | | RCP2.6 | RCP4.5 | RCP8.5 |
| (Simmons and Matthews 2016) | 2000-2100 | -0.35 / +0.42 | -0.29 / + 0.37 | -0.34 / + 0.35 |
| (Davies-Barnard et al. 2014) | 2005-2100 | -0.01 / +0.04 | +0.14 / -0.08 | -0.015 / +0.04 |
| (Boysen et al. 2014) | 2005-2100 | | | +0.04 / +0.08 0 / +0.05 +0.08 / +0.06 -0.20 / +0.13 -0.06 / +0.33 |

37
38 Two other projected land cover change scenarios have been examined (RCP4.5 and RCP2.6; Table 2.5 ;
39 Figure 2.14) but only one climate modelling experiment has been carried out for each, to estimate the
40 biophysical impacts on climate of those changes (Davies-Barnard et al. 2015). For RCP2.6, earth system and
41 dynamic global vegetation models agree on a systematic biogeochemical warming resulting from the
42 imposed land cover changes, ranging from +0.03 to +0.28°C (Brovkin et al. 2013a), which is significant
43 compared to the projected mean climate warming of +1°C by the end of the 21st century (ranging from 0.3 to
44 1.7°C depending on the models, compared to 1986-2005). A very small biophysical cooling is expected from
45 the one estimate. For RCP4.5 biophysical warming is expected from only one estimate, and results from a

1 projected large forestation in the temperate and high latitudes. There is no agreement on the sign of the
2 biogeochemical effect: there are as many studies predicting cooling as warming, whichever the method to
3 compute those effects (earth system models or dynamic global vegetation models).
4

5 Previous scenarios (Special Report on Emission Scenarios (SRES), results of climate studies using those
6 scenarios were reported in AR4) displayed larger land use changes than the more recent ones (RCP, AR5).
7 There is *low confidence* from some of those previous scenarios (SRES A2 and B1) of a small warming effect
8 (+0.2 to +0.3°C) of anthropogenic land cover change on mean global climate, this being dominated by the
9 release of CO₂ in the atmosphere from land conversions (Sitch et al. 2005). This additional warming remains
10 quite small when compared to the one resulting from the combined anthropogenic influences [+1.7°C for
11 SRES B1 and +2.7°C for SRES A2]. A global biophysical cooling of -0.14°C is estimated in response to the
12 extreme land cover change projected in SRES A2, a value that far exceeds the impacts of historical land use
13 changes (-0.05°C) calculated using the same climate model (Davin et al. 2007). The authors derived a
14 biophysical climatic sensitivity to land use change of about -0.3°C W.m⁻² for their model, whereas a warming
15 of about 1°C W.m⁻² is obtained in response to changes in atmospheric CO₂ concentration.
16

17 Those studies generally do not report on changes in atmospheric variables other than surface air temperature,
18 thereby limiting our ability to assess the effects of anthropogenic land cover changes on regional climate
19 Sitch et al. (2005). However, reported small reductions in rainfall via changes in biophysical properties of
20 the land, following the massive tropical deforestation in SRES A2 (+0.5 and +0.25 mm day⁻¹ respectively in
21 the Amazon and Central Africa). They also report opposite changes, that is increased rainfall of about 0.25
22 mm day⁻¹ across the entire tropics and subtropics, triggered by biogeochemical effects of this same
23 deforestation.
24

25 **2.5.1.2.2 At the regional level**

26 In regions that will undergo land cover changes, dampening of the future anthropogenic warming can be as
27 large as -26% while enhancement is always smaller than 9% within RCP8.5 by the end of the 21st century
28 (Boysen et al. 2014). Voltaire (2006) show that, by 2050 and following the SRES B2 scenario, the
29 contribution of land cover changes to the total temperature change can be as large as 15% in many boreal
30 regions, and as large as 40% in south western tropical Africa. Feddema et al. (2005) simulate large decreases
31 in the diurnal temperature range in the future (2050 and 2100 in SRES B1 and A2) following tropical
32 deforestation in both scenarios. In the Amazon for example the diurnal temperature range is lowered by
33 2.5°C due to increases in minimum temperature while little change is obtained for the maximum value.

34 There is thus *medium evidence* that future anthropogenic land cover change will have a significant effect on
35 regional temperature via biophysical effects in many regions of the world. There is however *no agreement*
36 on whether warming will be dampened or enhanced and there is *no agreement* on the sign of the contribution
37 across regions.
38

39 There are very few studies that go beyond analysing the changes in mean surface air temperature. Some
40 studies attempted to look at global changes in rainfall and found no significant influence of future land cover
41 changes (Brovkin et al. 2013a; Sitch et al. 2005; Feddema et al. 2005). Quesada et al. (2017a,b) however
42 carried out a systematic multi-model analysis of the response of a number of atmospheric, radiative and
43 hydrological variables (e.g. rainfall, sea level pressure, geopotential height, wind speed, soil-moisture,
44 turbulent heat fluxes, shortwave and longwave radiation, cloudiness) to RCP8.5 land cover scenario. In
45 particular, they found a significant reduction of rainfall in 6 out of 8 monsoon regions studied (Figure 2.16)
46 of about 1.9% to 3% (which means more than -0.5mm day⁻¹ in some areas) in response to future
47 anthropogenic land cover changes. Including those changes in global climate models reduces the projected
48 increase in rainfall by about 9% to 41% in those same regions, when all anthropogenic forcings are
49 accounted for (30% in the Global Monsoon region as defined by (Wang and Ding 2008)). In addition, they
50 found a shortening of the monsoon season of one to four days. They conclude that the projected future
51 increase in monsoon rains may be overestimated by those models that do not yet include biophysical effects
52 of land cover changes. Overall, the regional hydrological cycle was found to be substantially reduced and
53 wind speed significantly strengthened in response to regional deforestation within the tropics, with
54 magnitude comparable to projected changes with all forcings (Quesada et al. 2017b).
55

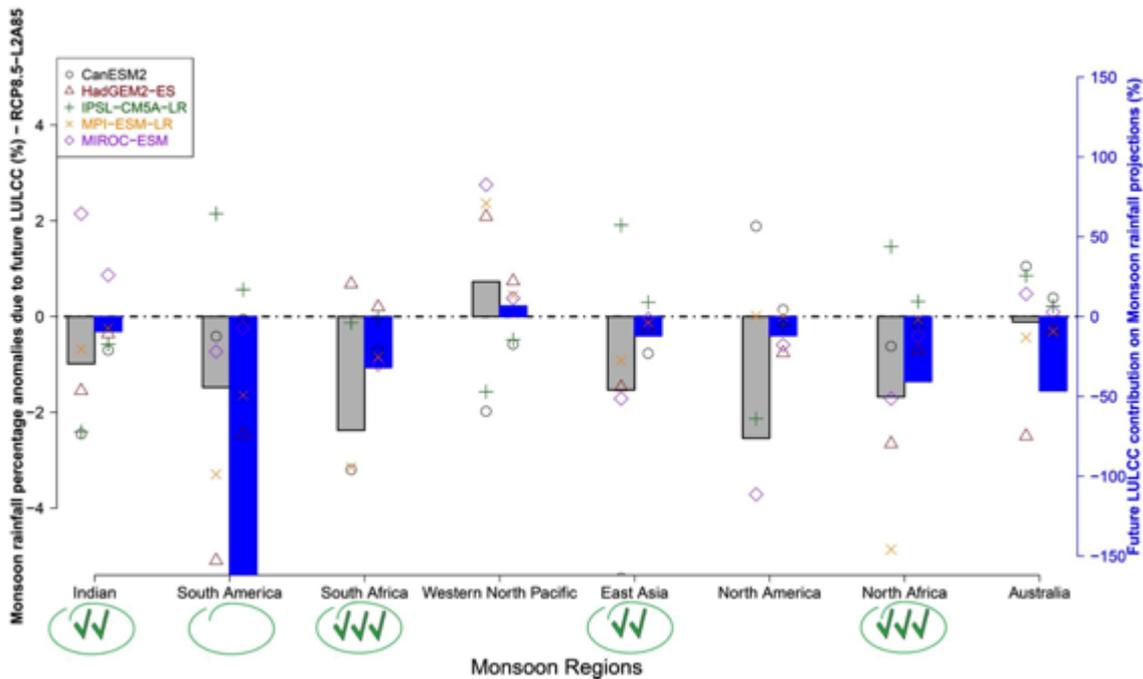


Figure 2.16 Changes in monsoon rainfall in RCP8.5 scenario resulting from projected changes in anthropogenic land cover, in eight monsoonal regions (%), grey bars). Differences are calculated between the end of the 21st century (2071–2100) and the end of the 20th century (1976–2005); percent change is calculated with reference to 1976–2005. Blue bars refer to the relative contribution of land cover changes (in %) to future rainfall projections: it is the ratio between the change in rainfall responding to land cover changes and the one responding to all anthropogenic changes (Quesada et al. 2017b). Negative values mean that changes in land cover have an opposite effect (dampening) on rainfall compared to the effects of all anthropogenic changes. Monsoon regions have been defined following (Yim et al. 2014). The changes have been simulated by five climate models (Brovkin et al. 2013, symbols). Results are shown for December-January-February for southern hemisphere regions, and for June-July-August for northern hemisphere regions. Statistical significance is given by green tick marks and circles: one, two, and three green tick marks are displayed for the regions where at least 80%, 75%, and 80th confidence level, respectively; green circles are added when the regional values are also significant at 90th confidence level. Note that future land cover change impacts on South American monsoon are neither significant nor robust among models, along with very small future projected changes in South American monsoon rainfall.

2.5.1.2.3 Effects on extremes

Results from a set of climate models have shown that the impact of future anthropogenic land cover change on extreme temperatures can be of similar magnitude as the changes arising from half a degree global mean annual surface temperature change (Hirsch et al. 2018). However, this study also found a lack of agreement between models with respect to the magnitude and sign of changes, thus making land cover change a factor of uncertainty in future climate projections.

2.5.2 Impacts of specific land use changes

2.5.2.1 Impacts of deforestation and forestation

Deforestation or forestation³, wherever it occurs, triggers simultaneously warming and cooling of the surface and of the atmosphere via changes in its various characteristics (Pitman 2003; Strengers et al. 2010; Bonan 2008b). Following deforestation, warming results from a) the release of CO₂ and other GHG in the atmosphere (biogeochemical impact) and subsequent increase in incoming infrared radiation at surface (greenhouse effect), b) a decreased in the total loss of energy through turbulent fluxes (latent and sensible

³ FOOTNOTE: The term « forestation » is used herein as this chapter does not distinguish between afforestation and reforestation. In model-based studies, simulations with and without trees are compared; in observation-based estimates, sites with and without trees are compared.

1 heat fluxes) resulting from reduced surface roughness, c) an increased incoming solar radiation following
2 reduced cloudiness that often (but not always) accompanies the decreased total evapotranspiration. Cooling
3 occurs in response to d) increased surface albedo that reduces the amount of absorbed solar radiation, e)
4 reduced incoming infrared radiation triggered by the decreased evapotranspiration and subsequent decrease
5 in atmospheric water vapour. Points b-c-d-e are referred to as biophysical effects. Deforestation and
6 forestation also alter rainfall and winds (horizontal as well as vertical as will be further discussed below).
7

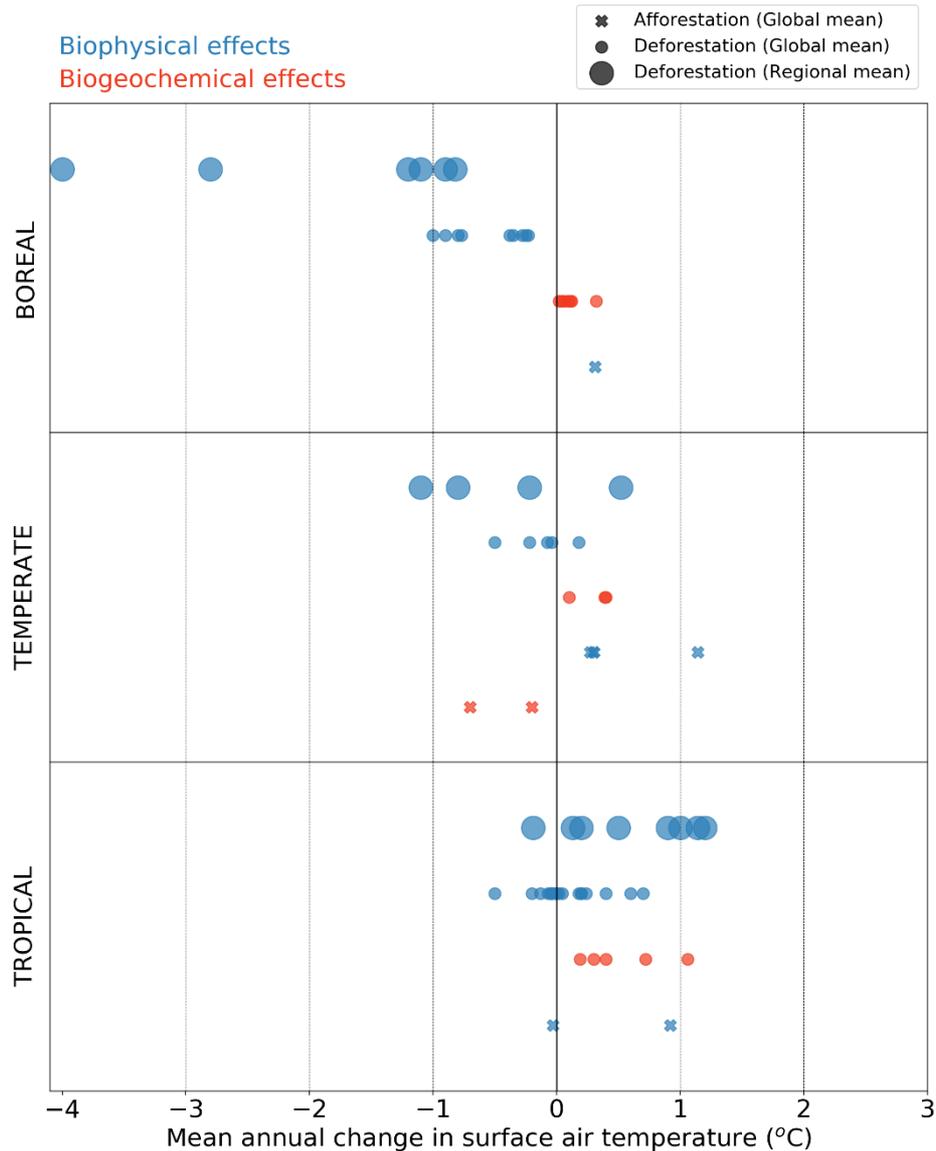
8 The literature that discusses the effects of forestation on climate is more limited than for deforestation, but
9 they reveal a similar climatic response with opposite sign as further discussed below. For each latitudinal
10 band (tropical, temperate and boreal) how very large scale deforestation or forestation impacts global mean
11 climate is examined, followed by examination of the large-scale changes in the specific latitudinal band and
12 end up with more regionally focused analysis. Large scale idealised deforestation or forestation experiments
13 are often carried out with global or regional climate models as they allow to understand and measure how
14 sensitive climate is to very large changes in land cover (similar to the instant doubling of CO₂ in climate
15 models to calculate the climatic sensitivity to GHGs). Details of the model-based studies discussed below
16 can be found in Table A2.2 in the Appendix.
17

18 2.5.2.1.1 *Global and regional impacts of deforestation/forestation in tropical regions*

19 A pan-tropical deforestation would lead to the net release of CO₂ from land and thus to mean global annual
20 warming, with model-based estimates of biogeochemical effects ranging from +0.19 to +1.06°C, with a
21 mean value of +0.53±0.32°C (Ganopolski et al. 2001; Snyder et al. 2004; Devaraju et al. 2015a; Longobardi
22 et al. 2016; Perugini et al. 2017). There is however *no agreement* between models on the magnitude and sign
23 of the biophysical effect of such changes at the global scale (the range spans from -0.5°C to +0.7°C with a
24 mean value of +0.1±0.27°C; Figure 2.17 ; e.g. (Devaraju et al. 2015c; Snyder 2010; Longobardi et al.
25 2016a)). This is the result of many compensation effects in action: increased surface albedo following
26 deforestation, decreased atmospheric water vapour content due to less tropical evapotranspiration, decreased
27 loss of energy from tropical land in the form of latent and sensible heat fluxes.
28

29 There is however *high confidence* that such large land cover change would lead to a mean biophysical
30 warming when averaged over the deforested land. A mean warming of +0.61±0.48°C is found over the entire
31 tropics. Reversely, biophysical regional cooling and global warming is expected from forestation (Wang et
32 al. 2014b; Bathiany et al. 2010a).
33

34 Large-scale deforestation (whether pan-tropical or imposed at the sub-continent level, e.g. the Amazon)
35 results in significant mean rainfall decrease (Lawrence and Vandecar 2015; Lejeune et al. 2015; Perugini et
36 al. 2017). In their review (Perugini et al. 2017) reported an average simulated decrease of $-288 \pm 75 \text{ mm yr}^{-1}$
37 (95%-confidence interval). Inversely large-scale forestation increases tropical rainfall by $41 \pm 21 \text{ mm yr}^{-1}$.
38 The magnitude of the change in precipitation strongly depends on the type of land cover conversion. For
39 instance, conversion of tropical forest to bare soil causes larger reductions in regional precipitation than
40 conversion to pasture (respectively $-470 \pm 60 \text{ mm yr}^{-1}$ and $-220 \pm 100 \text{ mm yr}^{-1}$). Biogeochemical effects in
41 response to pan-tropical deforestation, particularly CO₂ release, are generally not taken into account in those
42 studies but could intensify the hydrological cycle and thus precipitation (Kendra Gotangco Castillo and
43 Gurney 2013).
44



1
 2 **Figure 2.17 Changes in mean annual surface air temperature (°C) in response to idealised large scale**
 3 **deforestation (circles) or forestation (crosses), estimated from a range of studies (see Table A2.2 in the**
 4 **Appendix for detailed information and references to the studies). Temperature changes resulting from**
 5 **biophysical processes (e.g. changes in physical land surface characteristics such as albedo,**
 6 **evapotranspiration, and roughness length) are illustrated using blue symbols; temperature changes**
 7 **resulting from biogeochemical processes (e.g. changes in atmospheric CO₂ composition) use red**
 8 **symbols. Small blue and red circles, and crosses, are model-based estimates of changes in temperature**
 9 **averaged globally. Large circles are estimates averaged only over the latitudinal band where**
 10 **deforestation is imposed.**
 11

12 Specific model-based deforestation studies have been carried out for Africa (Hagos et al. 2014; Boone et al.
 13 2016; Xue et al. 2016; Nogherotto et al. 2013; Hartley et al. 2016; Klein et al. 2017; Abiodun et al. 2012),
 14 southern America (Butt et al. 2011; Wu et al. 2017; Spracklen and Garcia-Carreras 2015; Lejeune et al.
 15 2015), South-East Asia (Ma et al. 2013b; Werth and Avissar 2005; Mabuchi et al. 2005; Tölle et al. 2017).
 16 All found decreases in evapotranspiration following deforestation (*high agreement*), resulting in surface
 17 warming despite the competing effect from increased surface albedo (*high agreement*). Changes in thermal
 18 gradients between deforested and adjacent regions, between land and ocean, affect horizontal surface winds
 19 (*high agreement*) and thus modify the areas where rainfalls as discussed in Section 2.5.4. An increase in the
 20 land-sea thermal contrast has been found in many studies as surface friction is reduced by deforestation, thus

1 increasing the monsoon flow in Africa and South America (Wu et al. 2017).

2
3 Observation-based estimates all agree that deforestation increases local land-surface and ambient air
4 temperatures in the tropics, while forestation has the reverse effect (*very high confidence*; (Prevedello et al.
5 2019; Schultz et al. 2017; Li et al. 2015b; Alkama and Cescatti 2016)). There is *very high confidence* that
6 forests are cooler than any shorter vegetation (crops, grasses, bare soil) during daytime due to larger
7 transpiration rates, and there is *high confidence* that the amplitude of the diurnal cycle is smaller in the
8 presence of forests.

9
10 Large-scale forestation scenarios of West Africa (Abiodun et al. 2012), eastern China (Ma et al. 2013a) or
11 Saharan and Australian deserts (Ornstein et al. 2009; Kemena et al. 2017) all concluded that regional surface
12 cooling is simulated wherever trees are grown (-2.5°C in the Sahel and -1°C in the Savanna area of West
13 Africa, up to -8°C in western Sahara, -1.21°C over land in eastern China) while cooling of the ambient air is
14 smaller (-0.16°C). In the case of Savanna forestation this decrease entirely compensates the GHG induced
15 future warming ($+1^{\circ}\text{C}$ following the SRES A1B scenario). West African countries thus have the potential to
16 reduce, or even totally cancel at some places, the GHG-induced warming in the deforested regions (Abiodun
17 et al. 2012). However, this is compensated by enhanced warming in adjacent countries (non-local effect).

18 19 **2.5.2.1.2 Global and regional impacts of deforestation/forestation in temperate regions**

20 As for the tropics, model-based experiments show that large-scale temperate deforestation would induce a
21 small mean global annual warming through the net release of CO_2 into the atmosphere (ranging from $+0.10$
22 to $+0.40^{\circ}\text{C}$ with a mean value of $+0.20 \pm 0.13^{\circ}\text{C}$, Figure 2.17), whereas there is less agreement on the sign of
23 the mean global annual temperature change resulting from biophysical processes: estimates range from -
24 0.5°C to $+0.18^{\circ}\text{C}$ with a mean value of $-0.13 \pm 0.22^{\circ}\text{C}$. There is also *very low agreement* on the mean annual
25 temperature change in the temperate zone ($-0.4 \pm 0.62^{\circ}\text{C}$; (Phillips et al. 2007; Snyder et al. 2004b;
26 Longobardi et al. 2016a; Devaraju et al. 2015a, 2018b)). There is *medium agreement* on a global and
27 latitudinal biophysical warming in response to forestation (Figure 2.17 ; (Laguë and Swann 2016a; Swann et
28 al. 2012a; Gibbard et al. 2005; Wang et al. 2014b)), but this is based on a smaller number of studies.

29
30 The lack of agreement at the annual scale among the climate models is however masking *rising agreement*
31 regarding seasonal impacts of deforestation at those latitudes. There is *high agreement* that temperate
32 deforestation leads to summer warming and winter cooling (Bright et al. 2017; Zhao and Jackson 2014;
33 Gálos et al. 2011, 2013; Wickham et al. 2013; Ahlswede and Thomas 2017; Anderson-Teixeira et al. 2012;
34 Anderson et al. 2011; Chen et al. 2012; Strandberg and Kjellström 2018). The winter cooling is driven by the
35 increased surface albedo, amplified by the snow-albedo feedback. In some models and when deforestation is
36 simulated for very large areas, the cooling is further amplified by high latitude changes in sea-ice and snow
37 extent (polar amplification). Summer warming occurs because the latent and sensible heat fluxes, that take
38 energy out of the surface, diminish with the smaller roughness length and lower evapotranspiration
39 efficiency of low vegetation as compared to tree canopies (Davin and de Noblet-Ducoudre 2010; Anav et al.
40 2010). Conversely, there is *high agreement* that forestation in North America or in Europe cools surface
41 climate during summer time, especially in regions where water availability can support large
42 evapotranspiration rates. In temperate regions with water deficits, the simulated change in evapotranspiration
43 following forestation will be insignificant while the decreased surface albedo will favour surface warming.

44
45 Observation-based estimates confirm the existence of a seasonal pattern of response to deforestation, with
46 colder winters anytime there is snow on the ground and anywhere soils are brighter than trees, and warmer
47 summers (Schultz et al. 2017; Wickham et al. 2014; Juang et al. 2007; Tang et al. 2018; Peng et al. 2014;
48 Zhang et al. 2014b; Prevedello et al. 2019; Li et al. 2015b; Alkama and Cescatti 2016). In contrast
49 forestation induces cooler summers wherever trees have access to sufficient soil moisture to transpire. The
50 magnitude of the cooling depends on the wetness of the area of concern (Wickham et al. 2013) as well as on
51 the original and targeted species and varieties implicated in the vegetation conversion (Peng et al. 2014;
52 Juang et al. 2007).

53
54 There is also *high confidence* from observation-based estimates that mean annual daytime temperatures are
55 warmer following deforestation, while night time temperatures are cooler (Schultz et al. 2017; Wickham et
56 al. 2014; Juang et al. 2007; Tang et al. 2018; Prevedello et al. 2019; Peng et al. 2014; Zhang et al. 2014b; Li

1 et al. 2015b; Alkama and Cescatti 2016). Deforestation then increases the amplitude of diurnal temperature
2 variations while forestation reduces it (*high confidence*). Two main reasons have been put forward to explain
3 why nights are warmer in forested areas: their larger capacity to store heat, and the existence of a nocturnal
4 temperature inversion bringing warmer air from aloft.

5
6 In addition to those seasonal and diurnal fluctuations, (Lejeune et al. 2018) found systematic warming of the
7 hottest summer days following historical deforestation in the northern mid-latitudes, and this echoes
8 (Strandberg and Kjellström 2018) who argue that the August 2003 and July 2010 heat-waves could have
9 been largely mitigated if Europe had been largely forested.

10
11 In a combined modelling of large-scale forestation of western Europe and climate change scenario (SRES
12 A2) (Gálos et al. 2013) found a relatively small dampening potential of additional forest on ambient air
13 temperature at the end of the 21st century when compared to the beginning (the cooling resulting from land
14 cover changes is -0.5°C whereas the GHG-induced warming exceeds 2.5°C). Influence on rainfall was
15 however much larger and significant. Projected annual rainfall decreases following warming were cancelled
16 in Germany and significantly reduced in both France and Ukraine through forestation. In addition forestation
17 also decreased the number of warming-induced dry days, but increased the number of extreme precipitation
18 events.

19
20 The net impact of forestation, combining both biophysical and biogeochemical effects, has been tested in the
21 warmer world predicted by RCP 8.5 scenario (Sonntag et al. 2016, 2018). The cooling effect from the
22 addition of 8 Mkm^2 of forests following the land use RCP 4.5 scenario was too small (-0.27°C annually) to
23 dampen the RCP 8.5 warming. It however reached about -1°C in some temperate regions and -2.5°C in
24 boreal ones. This is accompanied by a reduction in the number of extremely warm days.

25 26 **2.5.2.1.3 Global and regional impacts of deforestation/forestation in boreal regions**

27 Consistent with what we have previously discussed for temperate and tropical regions, large-scale boreal
28 deforestation induces a biogeochemical warming of $+0.11 \pm 0.09^{\circ}\text{C}$ (Figure 2.17). But contrary to those other
29 latitudinal bands, the biophysical effect is a consistent cooling across all models ($-0.55 \pm 0.29^{\circ}\text{C}$ when
30 averaged globally). It is also significantly larger than the biogeochemical warming (e.g. (Dass et al. 2013;
31 Longobardi et al. 2016a; Devaraju et al. 2015a; Bathiany et al. 2010a; Devaraju et al. 2018b)). It is driven by
32 the increased albedo, enhanced by the snow-albedo feedback as well as by an increase in sea-ice extent in the
33 Arctic. Over the boreal lands, the cooling is as large as $-1.8 \pm 1.2^{\circ}\text{C}$. This mean annual cooling however
34 masks out a seasonal contrast as discussed in (Strandberg and Kjellström 2018) and (Gao et al. 2014): during
35 summer time, following the removal of forest, the decreased evapotranspiration results in a significant
36 summer warming that outweighs the effect of an increased albedo effect.

37
38 The same observation-based estimates as discussed in the previous sub-section show similar patterns as for
39 the temperate latitudes: seasonal and daily contrasts. (Schultz et al. 2017) however found that mean annual
40 nighttime changes are as large as daytime ones in those regions (mean annual nocturnal cooling $-1.4 \pm 0.10^{\circ}\text{C}$,
41 balanced by mean annual daytime warming of $1.4 \pm 0.04^{\circ}\text{C}$). This contrasts with both temperate and tropical
42 regions where daytime changes are always larger than nighttime ones.

43
44 Arora and Montenegro (2011) combined large-scale forestation and climate change scenario (SRES A2):
45 forestation of either 50% or 100% of the total agricultural area was gradually prescribed between years 2011
46 and 2060 everywhere. In addition, boreal, temperate and tropical forestation have been tested separately.
47 Both biophysical and biogeochemical effects were accounted for. The net simulated impact of forestation
48 was a cooling varying from -0.04°C to -0.45°C , depending on the location and magnitude of the additional
49 forest cover. It was, however, quite marginal compared to the large global warming resulting from
50 anthropogenic GHG emissions ($+3^{\circ}\text{C}$ at the end of the 21st century). In their experiment, forestation in
51 boreal regions led to biophysical warming and biogeochemical cooling that compensated each other, whereas
52 forestation in the tropics led to both biophysical and biogeochemical cooling. The authors concluded that
53 tropical forestation is three times more effective in cooling down climate than are boreal or temperate
54 forestation.

2.5.2.1.4 Conclusion

In conclusion, planting trees will always result in capturing more atmospheric CO₂ and thus in mean annual cooling of the globe (*very high confidence*). At the regional level however the magnitude and sign of the local temperature change depends on a) where forestation occurs, b) its magnitude, c) the level of warming under which the land cover change is applied and d) the land conversion type. This is because the background climatic conditions (e.g., precipitation and snow regimes, mean annual temperature) within which the land cover changes occur vary across regions (Pitman et al. 2011; Montenegro et al. 2009; Juang et al. 2007; Wickham et al. 2014; Hagos et al. 2014; Voltaire 2006; Feddema et al. 2005; Strandberg and Kjellström 2018). In addition there is *high confidence* that estimates of the influence of any land cover or land use change on surface temperature from the sole consideration of the albedo and the CO₂ effects is incorrect as changes in turbulent fluxes (i.e., latent and sensible heat fluxes) are large contributors to local temperature change (Bright et al. 2017).

There is *high confidence* that in boreal and temperate latitudes the presence of forest cools temperature in warmer locations and seasons provided that the soil is not dry, whereas it warms temperature in colder locations and seasons provided the soil is brighter than trees or covered with snow. In the humid tropics forestation increases evapotranspiration year round and thus decreases temperature (*high confidence*). In tropical areas with a strong seasonality of rainfall, forestation will also increase evapotranspiration year round unless the soil becomes too dry. In all regions there is *medium confidence* that the diurnal temperature range decreases with increasing forest cover, with potentially reduced extreme values of temperature

Although there is not enough literature yet that rigorously compares both biophysical and biogeochemical effects of realistic scenarios of forestation, there is *high confidence* that, at the local scale (that is where the forest change occurs) biophysical effects on surface temperature are far more important than the effects resulting from the changes in emitted CO₂.

What is lacking in the literature as of today is an estimate of the impacts natural disturbances in forests will have on local climates and on the build-up of atmospheric CO₂. (O'Halloran et al. 2012) for example illustrated with many examples that changes in albedo following disturbances can result in radiative forcing changes opposite to and as large as the ones resulting from the associated changes in the net release of CO₂ by land. The resulting climate effects depend on the duration of the perturbation and of the following recovery of vegetation.

2.5.2.2 Impacts of changes in land management

There have been little changes in net cropland area over the past 50 years (at the global scale) compared to continuous changes in land management (Erb et al. 2017). Similarly, in Europe change in forest management was a very significant anthropogenic land change. Management affects water, energy and GHG fluxes exchanged between the land and the atmosphere, and thus temperature and rainfall, sometimes to the same extent as changes in land cover do as discussed in (Luyssaert et al. 2014b).

The effects of irrigation, which is a practice that has been substantially studied, and one attempt to manage solar radiation via increases in cropland albedo (geoengineering the land) is assessed, along with discussion of recent findings on the effects of forest management on local climate, although there is not enough literature yet on this topic to carry out a real assessment. The effects of urbanisation on climate are assessed in a specific cross-chapter box within this chapter (Cross-Chapter Box 4 : Climate change and urbanisation, in this chapter).

There are a number of other practices that exist, some of them being reported in Section 2.6 and chapter 6 whose importance for climate mitigation has been examined. There is however not enough literature available for assessing their biophysical effect on climate. Few papers are generally found per agricultural practice, e.g., (Jeong et al. 2014b) for double cropping, (Bagley et al. 2017) for the timing of the growing season, and (Erb et al. 2017) for a review of ten management practices.

Similarly there are very few studies that have examined how choosing species varieties and harvesting strategies in forest management impacts climate through biophysical effects, and how those effects compare to the consequences of the chosen strategies on the net CO₂ sink of the managed forest. The modelling

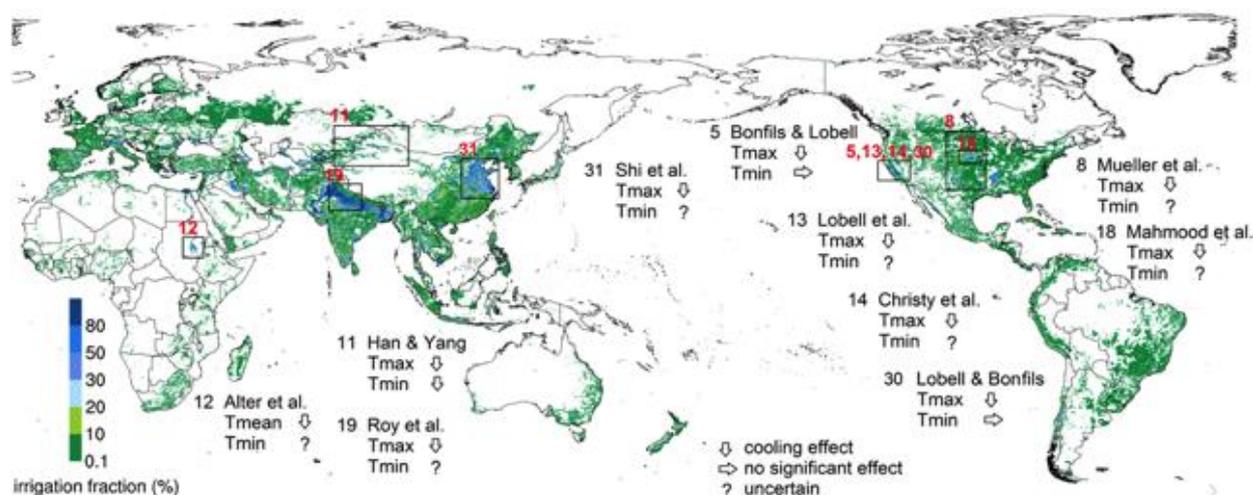
1 studies highlight the existence of competing effects between e.g. the capacity of certain species to store more
2 carbon than others (thus inducing cooling) while at the same time reducing the total evapotranspiration loss
3 and absorbing more solar radiation via lower albedo (thus inducing warming) (Naudts et al. 2016a; Luysaert
4 et al. 2018).

6 2.5.2.2.1 *Irrigation*

7 There is substantial literature on the effects of irrigation on local, regional and global climate as this is a
8 major land management. There is *very high confidence* that irrigation increases total evapotranspiration,
9 increases the total amount of water vapour in the atmosphere, and decreases mean surface daytime
10 temperature within the irrigated area and during the time of irrigation (Bonfils and Lobell 2007; Alter et al.
11 2015; Chen and Jeong 2018; Christy et al. 2006; Im and Eltahir 2014; Im et al. 2014; Mueller et al. 2015).
12 Decreases in maximum daytime temperature can locally be as large as -3°C to -8°C (Cook et al. 2015; Han
13 and Yang 2013; Huber et al. 2014; Alter et al. 2015; Im et al. 2014). Estimates of the contribution of
14 irrigation to past historical trends in ambient air temperature vary between -0.07°C and $-0.014^{\circ}\text{C}/\text{decade}$ in
15 Northern China (Han and Yang 2013; Chen and Jeong 2018) while being quite larger in California (-0.14°C
16 to $-0.25^{\circ}\text{C}/\text{decade}$; (Bonfils and Lobell 2007)). Surface cooling results from increased energy being taken up
17 from the land via larger evapotranspiration rates. In addition, there is growing evidence from modelling
18 studies that such cooling can locally mitigate the effect of heatwaves (Thiery et al. 2017; Mueller et al.
19 2015).

20
21 There is *no agreement* on changes in nighttime temperatures as discussed in (Chen and Jeong 2018) who
22 summarised the findings from observations in many regions of the World (India, China, North America and
23 eastern Africa; Figure 2.18). Where nighttime warming is found (Chen and Jeong 2018; Christy et al. 2006),
24 two explanations are put forward: the first is an increase in incoming long-wave radiation in response to
25 increased atmospheric water vapour content (greenhouse effect); the second is an increased storage of heat in
26 the soil during daytime, because of the larger heat capacity of a moister soil, heat that is then released to the
27 atmosphere at night.

28
29 There is *robust evidence* from modelling studies that implementing irrigation enhances rainfall although
30 there is *very low confidence* on where this increase occurs. When irrigation occurs in Sahelian Africa, during
31 the monsoon period, rainfall is decreased over the irrigated areas (*high agreement*) and increases south-west
32 if the crops are located in western Africa (Alter et al. 2015) and east / north-east when crops are located
33 further East in Sudan (Im and Eltahir 2014; Im et al. 2014) The cooler irrigated surfaces in the Sahel,
34 because of their greater evapotranspiration, inhibits convection and creates an anomalous descending motion
35 over crops that suppresses rainfall but influences the circulation of monsoon winds. Irrigation in India occurs
36 prior to the start of the monsoon season and the resulting land cooling decreases the land-sea temperature
37 contrast. This can delay the onset of the Indian monsoon and decrease its intensity (Niyogi et al. 2010;
38 Guimberteau et al. 2012). Results from (De Vrese et al. 2016a) modelling study suggest that part of the
39 excess rainfall triggered by Indian irrigation falls westward, in the horn of Africa. The theory behind those
40 local and downwind changes in rainfall support the findings from the models but we do not yet have
41 sufficient literature to robustly assess the magnitude and exact location of the expected changes driven by
42 irrigation.



1
2 **Figure 2.18** Global map of areas equipped for irrigation (colours), expressed as a percentage of total area, or
3 irrigation fraction (Siebert et al. 2013). Numbered boxes show regions where irrigation causes cooling (down
4 arrow) of surface mean (Tmean), maximum (Tmax) or minimum (Tmin) temperature, or else no significant
5 effect (right arrow) or where the effect is uncertain (question mark), based on observational studies as
6 reviewed in (Chen and Jeong 2018). Tmax refers to the warmest daily temperature while Tmin to the coldest
7 one which generally occurs at night. References are (Alter et al. 2015; Han and Yang 2013; Roy et al. 2007;
8 Shi et al. 2013; Bonfils and Lobell 2007; Lobell et al. 2008; Lobell and Bonfils 2008; Christy et al. 2006;
9 Mahmood et al. 2006; Mueller et al. 2015)

11 2.5.2.2.2 Cropland albedo

12 Various methods have been proposed to increase surface albedo in cropland and thus reduce locally surface
13 temperature (*high confidence*): choose ‘brighter’ crop varieties (Ridgwell et al. 2009; Crook et al. 2015;
14 Hirsch et al. 2017; Singarayer et al. 2009; Singarayer and Davies-Barnard 2012), abandon tillage (Lobell et
15 al. 2006; Davin et al. 2014), include cover crops into the rotation in areas where soils are darker than
16 vegetation (Carrer et al. 2018; Kaye and Quemada 2017) or use greenhouses (as in (Campra et al. 2008), see
17 (Seneviratne et al. 2018) for a review).

18 Whatever the solution chosen, the induced reduction in absorbed solar radiation cools the land, more
19 specifically during the hottest summer days ((Davin et al. 2014; Wilhelm et al. 2015); *low confidence*)
20 (Figure 2.19). Changes in temperature are essentially local and seasonal (limited to crop growth season) or
21 sub-seasonal (when resulting from inclusion of cover crop or tillage suppression). Such management action
22 on incoming solar radiation thus holds the potential to counteract warming in cultivated areas during crop
23 growing season.

24 Introducing cover crops into a rotation can also have a warming effect in areas where vegetation has a darker
25 albedo than soil, or in winter during snow periods if the cover crops or their residues are tall enough to
26 overtop the snow cover (Kaye and Quemada 2017; Lombardozzi et al. 2018). In addition evapotranspiration
27 greater than that of bare soil during this transitional period reduces soil temperature (Ceschia et al. 2017).
28 Such management strategy can have another substantial mitigation effect as it allows to store carbon in the
29 soil and to reduce both direct and indirect N₂O emissions (Basche et al. 2014; Kaye and Quemada 2017), in
30 particular if fertilisation of the subsequent crop is reduced (Constantin et al. 2010, 2011). The use of cover
31 crops thus improves substantially the GHG budget of croplands (Kaye and Quemada 2017; Tribouillois et al.
32 2018). More discussion on the role of the management practices for mitigation can be found in section 2.6
33 and chapter 6.

34 Only a handful of modelling studies have looked at effects other than changes in atmospheric temperature in
35 response to increased cropland albedo. (Seneviratne et al. 2018) have found significant changes in rainfall
36 following an idealised increase in cropland albedo, especially within the Asian monsoon regions. The
37 benefits of cooler temperature on production, resulting from increased albedo, is cancelled by decreases in
38 rainfall that are harmful for crop productivity. The rarity of a concomitant evaluation of albedo management
39 impact on crop productivity prevents us from providing a robust assessment of this practice in terms of both
40 climate mitigation and food security.

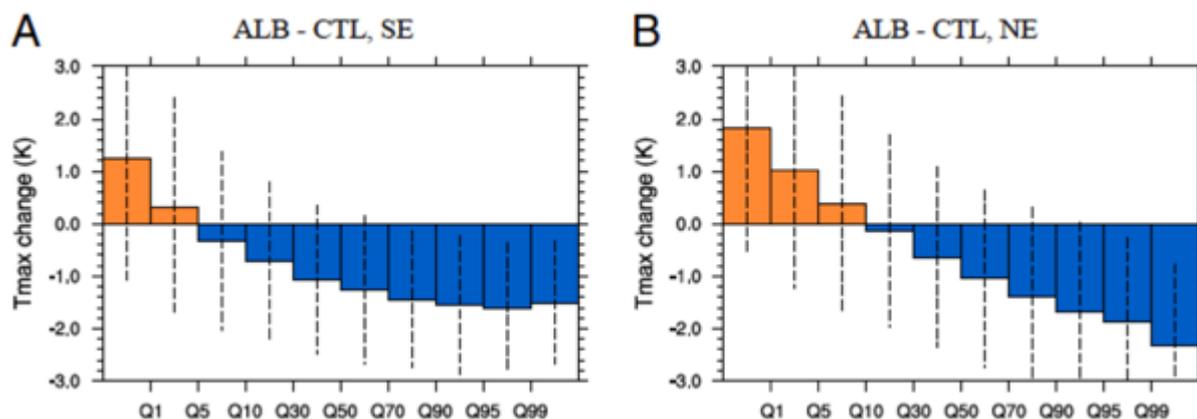
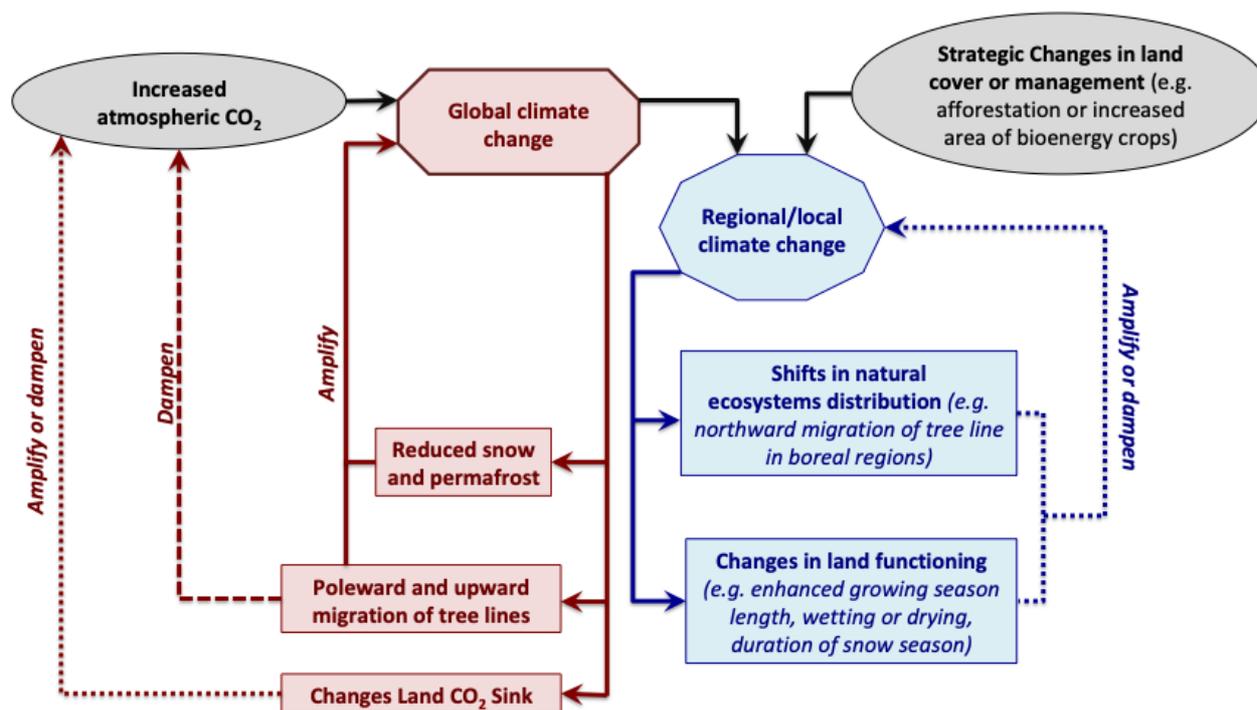


Figure 2.19 Change in summer (July-August) daily maximum temperature ($^{\circ}\text{C}$) resulting from increased surface albedo in unploughed versus ploughed land, in (A) Southern and (B) Northern Europe, during the period 1986–2009. Changes are simulated for different quantiles of the daily maximum temperature distribution, where Q1 represents the coolest 1% and Q99 the warmest 1% of summer days. Only grid cells with more than 60% of their area in cropland are included. The dashed bars represent the standard deviation calculated across all days and grid points. SE refers to southern Europe (below 45°N) and NE to northern Europe (above 45°N)

2.5.3 Amplifying / dampening climate changes via land responses

Section 2.1 and Box 2.1: illustrates the various mechanisms through which land can affect the atmosphere and thereby climate and weather. Section 2.2 illustrates the many impacts climate changes have on the functioning of land ecosystems. Section 2.3 discusses the effects future climatic conditions on the capacity of the land to absorb anthropogenic CO_2 , which then controls the sign of the feedback to the initial global warming. Sections 2.5.1 and 2.5.2 show effects of changes in anthropogenic land cover or land management on climate variables or processes. Land has thus the potential to dampen or amplify the GHG-induced global climate warming or can be used as a tool to mitigate regional climatic consequences of global warming such as extreme weather events, in addition to increasing the capacity of land to absorb CO_2 (Figure 2.20).

Land-to-climate feedbacks are difficult to assess with global or regional climate models as both types of models generally omit a large number of processes. Among these are 1) the response of vegetation to climate change in terms of growth, productivity, and geographical distribution, 2) the dynamics of major disturbances such as fires, 3) the nutrients dynamics, and 4) the dynamics and effects of short-lived chemical tracers such as biogenic volatile organic compounds (Section 2.4). Therefore, only those processes that are fully accounted for in climate models are considered here.

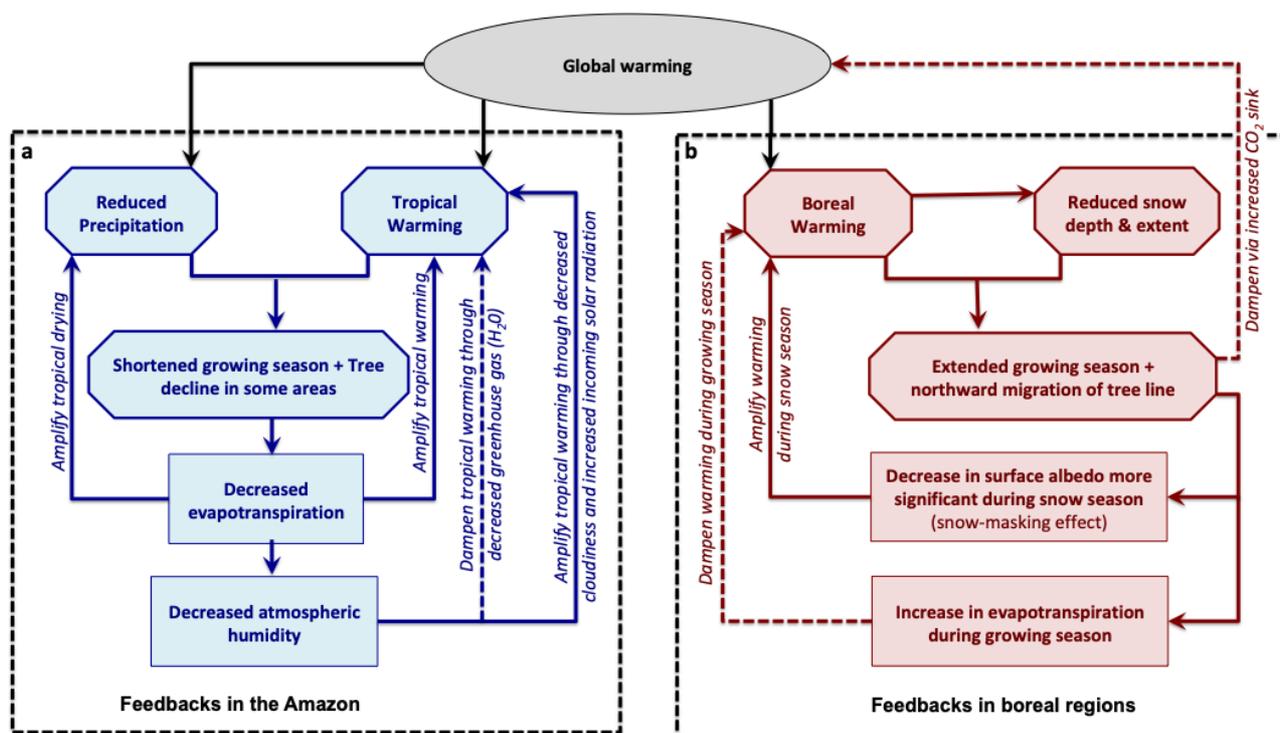


1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22

Figure 2.20 Schematics of the various ways land has been shown, in the literature, to either amplify or dampen the initial GHG-induced climatic change, at the global scale (left panel and red boxes and arrows) or at the regional/local levels (right panel and blue boxes and arrows). Grey arrows and boxes refer to what we consider herein as imposed changes, that is the initial atmospheric GHG content as well as anthropogenic land cover change and land management. Dampening feedbacks are represented with dashed lines, amplifying ones with solid lines and the feedbacks for which the direction may be variable are represented using dotted lines. The feedbacks initiated by changes in snow and permafrost areas in boreal regions is discussed in Section 2.5.3.2, the ones initiated by changes in ecosystem distribution are discussed in Sections 2.5.3.1, 2.5.1 and 2.5.2, and the feedbacks related to changes in the land functioning are discussed in Sections 2.5.3.3, 2.5.1, as well as 2.3 and 2.5 (for changes in net CO₂ fluxes). References supporting this figure can be found in each of those sections.

2.5.3.1 Effects of changes in land cover and productivity resulting from global warming

In boreal regions, the combined northward migration of the treeline and increased growing season length in response to increased temperatures in those regions (see Section 2.2) will have positive feedbacks both on global and regional annual warming (*high confidence*; Garnaud and Sushama 2015; Jeong et al. 2014a; O’ishi and Abe-Ouchi 2009; Port et al. 2012; Strengers et al. 2010). The warming resulting from the decreased surface albedo remains the dominant signal in all modelling studies at the annual time scale and during the snow season, while cooling is obtained during the growing season (see Section 2.5.2.1; Figure 2.21 right panel).



1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20
21
22
23
24
25
26
27
28
29
30
31
32
33
34
35
36

Figure 2.21 Schematic illustration of the processes through which the effects of global warming in the a) Amazon (left panel, blue boxes and arrows) and b) boreal regions (right panel, red boxes and arrows) feedback on the regional climate change. In boreal regions the sign of the feedbacks depends on the season, although annually global warming is further enhanced in those regions. Dashed lines illustrate negative feedbacks while solid lines indicate positive feedbacks. References supporting this figure can be found in the text.

In the tropics climate change will cause both greening and browning (see Section 2.2). Where global warming provokes decrease in rainfall, the induced decrease in biomass production leads to increased local warming (Port et al. 2012; Wu et al. 2016; Yu et al. 2016; *high confidence*). The reverse is true where warming generates increases in rainfall and thus greening. As an example, (Port et al. 2012) simulated decreases in tree cover and shortened growing season in the Amazon, despite the CO₂ fertilisation effects, in response to both future tropical warming and reduced precipitation (Figure 2.21, left panel). This browning of the land decreases both evapotranspiration and atmospheric humidity. The warming driven by the drop in evapotranspiration is enhanced via decreases in cloudiness that increases incoming solar radiation, and is dampened by reduced water vapour greenhouse radiation.

There is *very low confidence* on how feedbacks affect rainfall in the tropics where vegetation changes may occur, as the sign of the change in precipitation depends on where the greening occurs and on the season (as discussed in Section 2.5.2). There is however *high confidence* that increased vegetation growth in the southern Sahel increases African monsoon rains (Yu et al. 2016; Port et al. 2012; Wu et al. 2016). Confidence on the direction of such feedbacks is also based on a significant number of paleoclimate studies that analysed how vegetation dynamics helped maintain a northward position of the African monsoon during the Holocene time period (9 to 6 kyr BP) (de Noblet-Ducoudré et al. 2000; Rachmayani et al. 2015).

2.5.3.2 Feedbacks to climate from high-latitude land-surface changes

In high latitudes, snow albedo and permafrost carbon feedbacks are the most well-known and most important surface-related climate feedbacks because of their large-scale impacts.

In response to ongoing and projected decrease in seasonal snow cover (Derksen and Brown 2012; Brutel-Vuilmet et al. 2013) warming is and will continue to be enhanced in boreal regions (*high confidence*; Brutel-Vuilmet et al. 2013; Perket et al. 2014; Thackeray and Fletcher 2015; Mudryk et al. 2017). One reason for this is the large reflectivity (albedo) the snow exerts on shortwave radiative forcing: the all-sky global land snow shortwave radiative effect is evaluated to be around $-2.5 \pm 0.5 \text{ W m}^{-2}$ (Flanner et al. 2011; Singh et al. 2015). In the Southern Hemisphere, perennial snow on the Antarctic is the dominant contribution, while in the Northern Hemisphere, this is essentially attributable to seasonal snow with a smaller contribution from

1 snow on glaciated areas. Another reason is the sensitivity of snow cover to temperature: (Mudryk et al. 2017)
2 recently showed that in the high latitudes, climate models tend to correctly represent this sensitivity, while in
3 mid-latitude and alpine regions, the simulated snow cover sensitivity to temperature variations tends to be
4 biased low. In total, the global snow albedo feedback is about $0.1 \text{ W m}^{-2} \text{ K}^{-1}$, which amounts to about 7% of
5 the strength of the globally dominant water vapour feedback (e.g., (Thackeray and Fletcher 2015)). While
6 climate models do represent this feedback, a persistent spread in the modelled feedback strength has been
7 noticed (Qu and Hall 2014) and, on average, the simulated snow albedo feedback strength tends to be
8 somewhat weaker than in reality (Flanner et al. 2011; Thackeray and Fletcher 2015) (*medium confidence*).
9 Various reasons for the spread and biases of the simulated snow albedo feedback have been identified,
10 notably inadequate representations of vegetation masking of snow in forested areas (Lorantý et al. 2014;
11 Wang et al. 2016c; Thackeray and Fletcher 2015).

12
13 The second most important potential feedback from land to climate relates to permafrost decay. There is *high*
14 *confidence* that, following permafrost decay from a warming climate, the resulting emissions of carbon
15 dioxide and/or methane (caused by the decomposition of organic matter in previously frozen soil) will
16 produce additional GHG-induced warming. There is however substantial uncertainty on the magnitude of
17 this feedback, although recent years have seen large progress in its quantification. Lack of agreement results
18 from several critical factors that carry large uncertainties. The most important are a) the size of the
19 permafrost carbon pool, b) its decomposability, c) the magnitude, timing and pathway of future high-latitude
20 climate change and d) the correct identification and model representation of the processes at play (Schuur et
21 al. 2015b). The most recent comprehensive estimates establish a total soil organic carbon storage in
22 permafrost of about $1500 \pm 200 \text{ Pg C}$ (Hugelius et al. 2014, 2013; Olefeldt et al. 2016), which is about 300
23 Pg C lower than previous estimates (*low confidence*). Important progress has been made in recent years at
24 incorporating permafrost-related processes in complex Earth System Models (e.g., (McGuire et al. 2018)),
25 but representations of some critical processes such as thermokarst formation are still in their infancy (Schuur
26 et al. 2015b). Recent model-based estimates of future permafrost carbon release (Koven et al. 2015; McGuire
27 et al. 2018) have converged on an important insight. Their results suggest that substantial net carbon release
28 of the coupled vegetation-permafrost system will probably not occur before about 2100 because carbon
29 uptake by increased vegetation growth will initially compensate for GHG releases from permafrost (*limited*
30 *evidence, high agreement*).

31 **2.5.3.3 Feedbacks related to changes in soil moisture resulting from global warming**

32 There is *medium evidence* but *high agreement* that soil moisture conditions influence the frequency and
33 magnitude of extremes such as drought and heat waves. Observational evidence indicates that dry soil
34 moisture conditions favour heat-waves, in particular in regions where evapotranspiration is limited by
35 moisture availability (Mueller and Seneviratne 2012; Quesada et al. 2012; Miralles et al. 2018; Geirinhas et
36 al. 2018; Miralles et al. 2014; Chiang et al. 2018; Dong and Crow 2019; Hirschi et al. 2014).

37
38
39 In future climate projections, soil moisture plays an important role in the projected amplification of extreme
40 heat-waves and drought in many regions of the world (*medium confidence*; (Seneviratne et al. 2013; Vogel et
41 al. 2017; Donat et al. 2018; Miralles et al. 2018)). In addition, the areas where soil moisture affects heat
42 extremes will not be located exactly where they are today. Changes in rainfall, temperature and thus
43 evapotranspiration will induce changes in soil moisture and therefore of where temperature and latent heat
44 flux will be negatively coupled (Seneviratne et al. 2006; Fischer et al. 2012). Quantitative estimates of the
45 actual role of soil moisture feedbacks are however very uncertain due to the *low confidence* in projected soil
46 moisture changes (IPCC 2013a), to weaknesses in the representation of soil moisture-atmosphere
47 interactions in climate models (Sippel et al. 2017; Ukkola et al. 2018; Donat et al. 2018; Miralles et al. 2018)
48 and to methodological uncertainties associated with the soil moisture prescription framework commonly
49 used to disentangle the effect of soil moisture on changes in temperature extremes (Hauser et al. 2017).

50
51 Where soil moisture is predicted to decrease in response to climate change in the subtropics and temperate
52 latitudes, this drying could be enhanced by the existence of soil moisture feedbacks (*low confidence* (Berg et
53 al. 2016)). The initial decrease in precipitation and increase in potential evapotranspiration and latent heat
54 flux, in response to global climate change, leads to decreased soil moisture at those latitudes and can
55 potentially amplify both. Such a feature is consistent with evidence that in a warmer climate land and
56 atmosphere will be more strongly coupled via both the water and the energy cycles (Dirmeyer et al. 2014;

1 Guo et al. 2006). This increased sensitivity of atmospheric response to land perturbations implies that
 2 changes in land uses and cover are expected, in the future, to have more impact on climate in the future than
 3 they do today.

4
 5 Beyond temperature, it has been suggested that soil moisture feedbacks influence precipitation occurrence
 6 and intensity. But the importance and even the sign of this feedback is still largely uncertain and debated
 7 (Tuttle and Salvucci 2016; Yang et al. 2018; Froidevaux et al. 2014; Guillod et al. 2015).

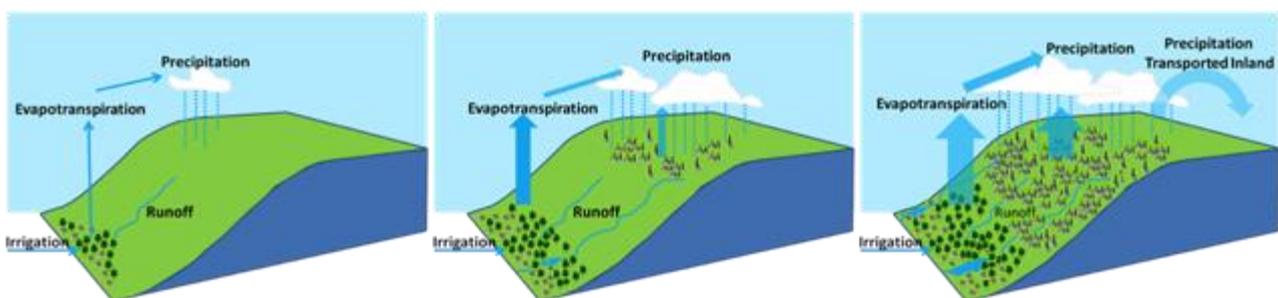
8 9 **2.5.4 Non-local and downwind effects resulting from changes in land cover**

10
 11 Changes in land cover or land management do not just have local consequences but also affect adjacent or
 12 more remote areas. Those non-local impacts may occur in three different ways.

13
 14 (1) Any action on land that affects photosynthesis and respiration has an impact on the atmospheric CO₂
 15 content as this GHG is well mixed in the atmosphere. This change in turn affects the downwelling long-wave
 16 radiation everywhere on the planet and contributes to global climate change. This is more thoroughly
 17 discussed in Section 2.6 where various land-based mitigation solutions are examined. Local land use changes
 18 thus have the potential to affect global climate via changes in atmospheric CO₂.

19
 20 (2) Any change in land cover or land management may impact local surface air temperature and moisture
 21 and thus sea-level pressure. Thermal, moisture and surface pressure gradients between the area of change
 22 and neighbouring areas are then modified and affect the amount of heat, water vapour and pollutants flowing
 23 out (downwind) of the area (e.g. Ma et al. 2013b; McLeod et al. 2017; Abiodun et al. 2012; Keys 2012).

24 Forests for example provide water vapour to the atmosphere which supports terrestrial precipitation
 25 downwind (Ellison et al. 2017a; Layton and Ellison 2016; Spracklen et al. 2012, 2018). Within a few days
 26 water vapour can travel several hundreds of kilometres before being condensed into rain and potentially
 27 being transpired again (Makarieva et al. 2009). This cascading moisture recycling (succession of
 28 evapotranspiration, water vapour transport and condensation-rainfall) has been observed in south America
 29 (Spracklen et al. 2018; Zemp et al. 2014; Staal et al. 2018; Spracklen et al. 2012). Deforestation can thus
 30 potentially decrease rainfall downwind, while combining 'small-scale' forestation and irrigation in the semi-
 31 arid region is susceptible to boost the precipitation-recycling mechanism with better vegetation growth
 32 downwind (Figure 2.22; (Ellison et al. 2017a; Layton and Ellison 2016)).
 33



34
 35 **Figure 2.22 Schematic illustration of how combined forestation and irrigation can influence downwind**
 36 **precipitation on mountainous areas (here in Los Angeles, California area), favour vegetation growth and**
 37 **feeds back to the forested area via increased runoff (Layton and Ellison 2016). Areas of forests**
 38 **plantation and irrigation are located on the left panel, whereas consequent downwind effects and**
 39 **feedbacks are illustrated in the middle and right panels.**
 40

41 (3) Many studies using global climate models have reported that the climatic changes resulting from changes
 42 in land are not limited to the lower part of the atmosphere but can reach the upper levels via changes in large
 43 scale ascent (convection) or descent (subsidence) of air. This coupling to the upper atmosphere triggers
 44 perturbations in large-scale atmospheric transport (of heat, energy and water) and subsequent changes in
 45 temperature and rainfall in regions located quite far away from the original perturbation (Figure 2.23, Laguë
 46 and Swann 2016; Feddema et al. 2005, badger & dirmeyer 2016, Garcia 2016, Stark 2015, Devaraju 2018,
 47 Quesada et al. (2017a)).
 48

49 De Vrese et al. (2016) for example, using a global climate model, found that irrigation in India could affect

1 regions as remote as eastern Africa through changes in the atmospheric transport of water vapour. At the
 2 onset of boreal spring (February to March) evapotranspiration is already large over irrigated crops and the
 3 resulting excess moisture in the atmosphere is transported south-westward by the low-level winds. This
 4 results in increases in precipitation as large as 1mm d⁻¹ in the horn of Africa. Such finding implies that if
 5 irrigation is to decrease in India, rainfall can decrease in eastern Africa where the consequences of drought
 6 are already disastrous.
 7



8
 9
 10 **Figure 2.23 Extra-tropical effects on precipitation due to deforestation in each of the three major tropical**
 11 **regions. Increasing (circles) and decreasing (triangles) precipitation result from complete deforestation of either**
 12 **Amazonia (red), Africa (yellow), or Southeast Asia (blue) as reviewed by (Lawrence and Vandecar 2015). Boxes**
 13 **indicate the area in which tropical forest was removed in each region. Numbers refer to the study from which**
 14 **the data were derived. Cited papers are the following (Avissar and Werth 2005; Gedney and Valdes 2000;**
 15 **Semazzi and Song 2001; Werth 2002; Mabuchi et al. 2005; Werth 2005)**
 16

17 Changes in sea-surface temperature have also been simulated in response to large-scale vegetation changes
 18 (Cowling et al. 2009; Davin and de Noblet-Ducoudre 2010; Wang et al. 2014b, Notaro Liu 2007).
 19 Most of those modelling studies have been carried out with land cover changes that are extremely large and
 20 often exaggerated with respect to reality. The existence of such teleconnections can thus be biased as
 21 discussed in Lorenz et al. (2016).
 22

23 In conclusion, there is *high confidence* that any action on land (for example to dampen global warming
 24 effects), wherever they occur, will not only have effects on local climate but also generate atmospheric
 25 changes in neighbouring regions, and potentially as far as few hundreds of kilometres downwind. More
 26 remote teleconnections, thousands of kilometres away from the initial perturbation, are impossible to observe
 27 and have only been reported by modelling studies using extreme land cover changes. There is *very low*
 28 *confidence* that detectable changes due to such long-range processes can occur.
 29
 30
 31

32 **Cross-Chapte Box 4: Climate Change and Urbanisation**

33
 34 Nathalie de Noblet-Ducoudré (France), Peng Cai (China), Sarah Connors (France/United Kingdom), Martin
 35 Dallimer (United Kingdom), Jason Evans (Australia), Rafiq Hamdi (Belgium), Gensuo Jia (China), Kaoru
 36 Kitajima (Japan), Christopher Lennard (South Africa), Shuaib Lwasa (Uganda), Carlos Fernando Mena
 37 (Ecuador), Soojeong Myeong (The Republic of Korea), Lennart Olsson (Sweden), Prajal Pradhan
 38 (Nepal/Germany), Lindsay Stringer (United Kingdom)

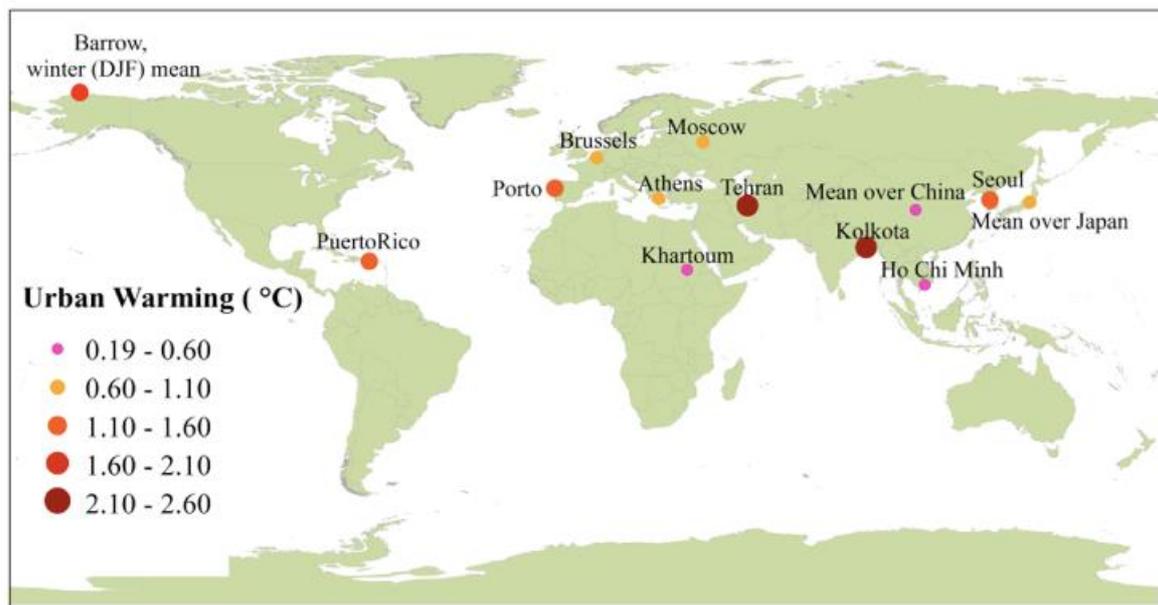
Cities extent, population, and expected growth

Despite only covering 0.4-0.9% of the global land surface (Esch et al. 2017; Zhou et al. 2015), over half the world's population live in towns and cities (United Nations 2017) generating around three-quarters of the global total carbon emissions from energy use (Creutzig et al. 2015b; Intergovernmental Panel on Climate Change 2014). Urban food consumption is a large source of these anthropogenic greenhouse gas emissions (Goldstein et al. 2017). In developed countries, per capita emissions are larger in small cities than bigger ones, while the opposite is found in developing countries (Gudipudi et al. 2019). Climate change is expected to increase the energy demand of people living in urban areas (Santamouris et al. 2015; Wenz et al. 2017).

In addition to being a driver of emissions, urbanisation contributes to forest degradation, converts neighbouring agricultural, forested, or otherwise undeveloped land to urban use, altering natural or semi-natural ecosystems both within and outside of urban areas (Du and Huang 2017). It has been identified as a major driver of land degradation as illustrated in Chapters 3, 4 and 5. Highly productive lands are experiencing the highest rate of conversion to urbanised landscapes (Nizeyimana et al. 2001; Pandey et al. 2018), affecting food security. Loss of agricultural land, and increased pollution and waste are some of key challenges arising from urbanisation and urban growth (Chen 2007). The proportion of urban population is predicted to reach ~70% by the middle of the century (United Nations 2017) with growth especially taking place in the developing world (Angel et al. 2011; Dahiya 2012). Urban sprawl is projected to consume 1.8–2.4% and 5% of the current cultivated land by 2030 and 2050 respectively (Pradhan et al. 2014; Brend'Amour et al. 2016) driven by both general population increase and migration from rural areas (Adger et al. 2015; Seto et al. 2011; Geddes et al. 2012). New city dwellers in developing countries will require land for housing to be converted from non-urban to urban land (Barbero-Sierra et al. 2013), indicating future degradation. These growing urban areas will experience direct and indirect climate change impacts, such as sea level rise and storm surges (Boettle et al. 2016; Revi et al. 2014), increasing soil salinity, and landslides from precipitation extremes. Furthermore, poorly planned urbanisation can increase people's risk to climate hazards as informal settlements and poorly built infrastructure are often the most exposed to hazards from fire, flooding, and landslides (Adger et al. 2015; Geddes et al. 2012; Revi et al. 2014). Currently, avoiding land degradation and maintaining/enhancing ecosystem services are rarely considered in planning processes (Kuang et al. 2017).

Climate change, urban heat island and threats specific to urban populations

Cities alter the local atmospheric conditions as well as those of the surrounding areas (Wang et al. 2016b; Zhong et al. 2017). There is *high confidence* that urbanisation increases mean annual surface air temperature in cities and in their surroundings, with increases ranging from 0.19°C to 2.60°C (Cross Chapter Box 4 Figure 1) (Torres-Valcárcel et al. 2015; Li et al. 2018a; Doan et al. 2016). This phenomenon is referred to as the urban heat island (UHI) effect (Oke et al. 2017; Bader et al. 2018). The magnitude and diurnal amplitude of the UHI varies from one city to another and depends on the local background climate (Wienert and Kuttler 2005; Zhao et al. 2014; Ward et al. 2016). There is nevertheless *high confidence* that urbanisation affects night time temperatures more substantially than daytime ones (Argüeso et al. 2014; Alghamdi and Moore 2015; Alizadeh-Choobari et al. 2016; Fujibe 2009; Hausfather et al. 2013; Liao et al. 2017; Sachindra et al. 2016; Camilloni and Barrucand 2012; Wang et al. 2017a; Hamdi 2010; Arsiso et al. 2018; Elagib 2011; Lokoshchenko 2017; Robaa 2013). In addition there is *high confidence* that the UHI effect makes heatwaves more intense in cities by 1.22°C to 4°C, particularly at night (Li and Bou-Zeid 2013; Li et al. 2017b; Hamdi et al. 2016; Founda and Santamouris 2017; Wang et al. 2017a). As there is a well-established relationship between extremely high temperatures and morbidity, mortality (Watts et al. 2015) and labour productivity (Costa et al. 2016), expected increase in extreme heat events with future climate change will worsen the conditions in cities.



Cross Chapter Box 4, Figure 1: Change in annual mean surface air temperature resulting from urbanisation (°C). Colour and size of the circles refer to the magnitude of the change. This map has been compiled using the following studies: (Kim et al. 2016; Sun et al. 2016; Chen et al. 2016a; Founda et al. 2015; Rafael et al. 2017; Hinkel and Nelson 2007; Chrysanthou et al. 2014; Dou et al. 2014; Zhou et al. 2016, 2017; Polydoros et al. 2018; Li et al. 2018a; Bader et al. 2018; Alizadeh-Choobari et al. 2016; Fujibe 2009; Lokoshchenko 2017; Torres-Valcárcel et al. 2015; Doan et al. 2016; Elagib 2011; Liao et al. 2017).

Individual city case studies show that precipitation mean and extremes are increased over and downwind of urban areas, especially in the afternoon and early evening when convective rise of the atmosphere is the strongest (*medium confidence*). The case studies covered: different inland and coastal US cities (M. et al. 2014; McLeod et al. 2017; Ganeshan and Murtugudde 2015); Dutch coastal cities (Daniels et al. 2016); Hamburg (Schlünzen et al. 2010); Shanghai (Liang and Ding 2017); Beijing (Dou et al. 2014); and Jakarta and Kuala Lumpur (Lorenz et al. 2016). Increased aerosol concentrations however can interrupt the precipitation formation process and thereby reduce heavy rainfall (Daniels et al. 2016; Zhong et al. 2017). Urban areas also experience altered water cycle in other aspects, the evaporative demand for plants in cities are increased by as much as 10% (Zipper et al. 2017) while high proportion of paving in cities mean that surface runoff of water is high (Hamdi et al. 2011; Pataki et al. 2011). In addition, water retention is lower in degraded, sealed soils beneath urban surfaces compared to intact soils. Increased surface water runoff, especially when and where the rainfall intensity is likely to intensify (IPCC 2013b), leads to a greater likelihood of flooding in urban areas without implementation of adaptation measures (Shade and Kremer 2019; Wang et al. 2013; Environmental Protection Agency 2015).

Urbanisation alters the stock size of soil organic carbon (SOC) and its stability. The conversion of vegetated land to urban land results in a loss of carbon stored in plants, while stresses associated with the urban environment (e.g., heat, limited water availability and pollution) reduce plant growth and survival in cities (Xu et al. 2016b). Overall, carbon densities or stocks decrease from natural land areas to the urban core along the rural-urban gradient (Tao et al. 2015; Zhang et al. 2015). For example the Seoul Forest Park, an urban park, shows a tenfold difference in SOC stocks across its land cover types (Bae and Ryu 2015). In Changchun in Northeast China, however, SOC density is higher in recreational forests within urban areas compared to a production forest (Zhang et al. 2015).

Urban air pollution as an environmental risk increases with climate change. Increased air temperatures can lead to reduced air quality by enhancing the formation of photochemical oxidants and increasing the concentration of air pollutants such as ozone, with corresponding threats to human health (Sharma et al. 2013). The occurrence of bronchial asthma and allergic respiratory diseases is increasing worldwide, and urban residents are experiencing poor air quality conditions more frequently than rural residents (D'Amato et al. 2010). Excess morbidity and mortality related to extremely poor air quality are found in many cities

1 worldwide (Harlan and Ruddell 2011). Some emissions that lead to reduced air quality are also contributors
2 to climate change (Shindell et al. 2018; de Coninck et al. 2018).
3

4 **Urban response options for climate change, desertification, land degradation and food security**

5 Urban green infrastructure (UGI; see glossary) has been proposed as a solution to mitigate climate change
6 directly through carbon sequestration (Davies et al. 2011; Edmondson et al. 2014). However, compared to
7 overall carbon emissions from cities, its mitigation effects are likely to be small (*medium confidence*). UGI
8 nevertheless has an important role in adapting cities to climate change (Demuzere et al. 2014; Sussams et al.
9 2015; Martin et al. 2016; Gill et al. 2007; Revi et al. 2014). Adaptation through UGIs is achieved through,
10 for example: (i) reduction in air temperature (Cavan et al. 2014; Di Leo et al. 2016; Feyisa et al. 2014; Zölch
11 et al. 2016; Li et al. 2019) which can help improve human health and comfort (e.g. (Brown and Nicholls
12 2015; Klemm et al. 2015)); (ii) reduction in the energy demands of buildings through the use of green roofs
13 and walls (e.g.(Coma et al. 2017)); (iii) reduction in surface water runoff and flood risk (Zeileňáková et al.
14 2017). Given that UGI necessarily involves the retention and management of non-sealed surfaces, co-
15 benefits for land degradation will also be apparent (Murata and Kawai 2018; Scalenghe and Marsan 2009)
16 (*limited evidence, high agreement*).
17

18 Urban agriculture is one aspect of UGI that has the potential to both meet some of the food needs of cities
19 and reduce land degradation pressures in rural areas (*low confidence*; e.g. Wilhelm and Smith 2018). Urban
20 agriculture has many forms, such as backyard gardening, allotments, plants on roof-tops or balconies, urban-
21 fringe/peri-urban agriculture, hydroponics, aquaponics, livestock grazing in open spaces and vertical farming
22 (Gerster-Bentaya 2013) (see also Section 5.6.5).
23

24 Consuming locally produced food and enhancing the efficiency of food processing and transportation can
25 minimise food losses, contribute to food security, and in some circumstances reduce GHG emissions (Brodt
26 et al. 2013; Michalský and Hooda 2015; Tobarra et al. 2018) (see also Section 5.5.2.3). Furthermore, urban
27 agriculture has the potential to counteract the separation of urban populations from food production. This
28 separation is one driver of the transition towards more homogeneous, high protein diets, which are associated
29 with increased greenhouse gas emissions (Goldstein et al. 2017; Moragues-Faus and Marceau 2018;
30 Magarini and Calori 2015). Barriers to the uptake of urban agriculture as a climate change mitigation option
31 include the need for efficient distribution systems to ensure lowered carbon emissions (Newman et al. 2012)
32 and the concern that urban agriculture may harbour pathogenic diseases, or that its products be contaminated
33 by soil or air pollution (Hamilton et al. 2014; Ercilla-Montserrat et al. 2018).
34

35 **In summary**

36 Climate change is already affecting the health and energy demand of large numbers of people living in urban
37 areas (*high confidence*; see also Section 2.2). Future changes to both climate and urbanisation will enhance
38 warming in cities and their surroundings, especially during heat waves (*high confidence*). Urban and peri-
39 urban agriculture, and more generally the implementation of urban green infrastructure, can contribute to
40 climate change mitigation (*medium confidence*) as well as to adaptation (*high confidence*), including co-
41 benefits for food security and reduced soil-water-air pollution.
42
43

2.6 Climate consequences of response options

Response options can affect climate mitigation and adaptation simultaneously, therefore this Special Report on Climate Change and Land (SRCCL) discusses land-based response options in an integrated way (Chapter 1). In this chapter we assess response options that have an effect on climate. A description of the full set of response options across the SRCCL can be found in Chapter 6, including the interplay between mitigation, adaptation, desertification, land degradation, food security and other co-benefits and trade-offs. Response options specific to desertification, degradation and food security are described in more detail in Chapters 3, 4 and 5.

Some response options lead to land use change and can compete with other land uses, including other response options, while others may free-up land that can be used for further mitigation/adaptation by reducing demand for land or products e.g. agricultural intensification, diet shifts, and reduction of waste (*high confidence*).

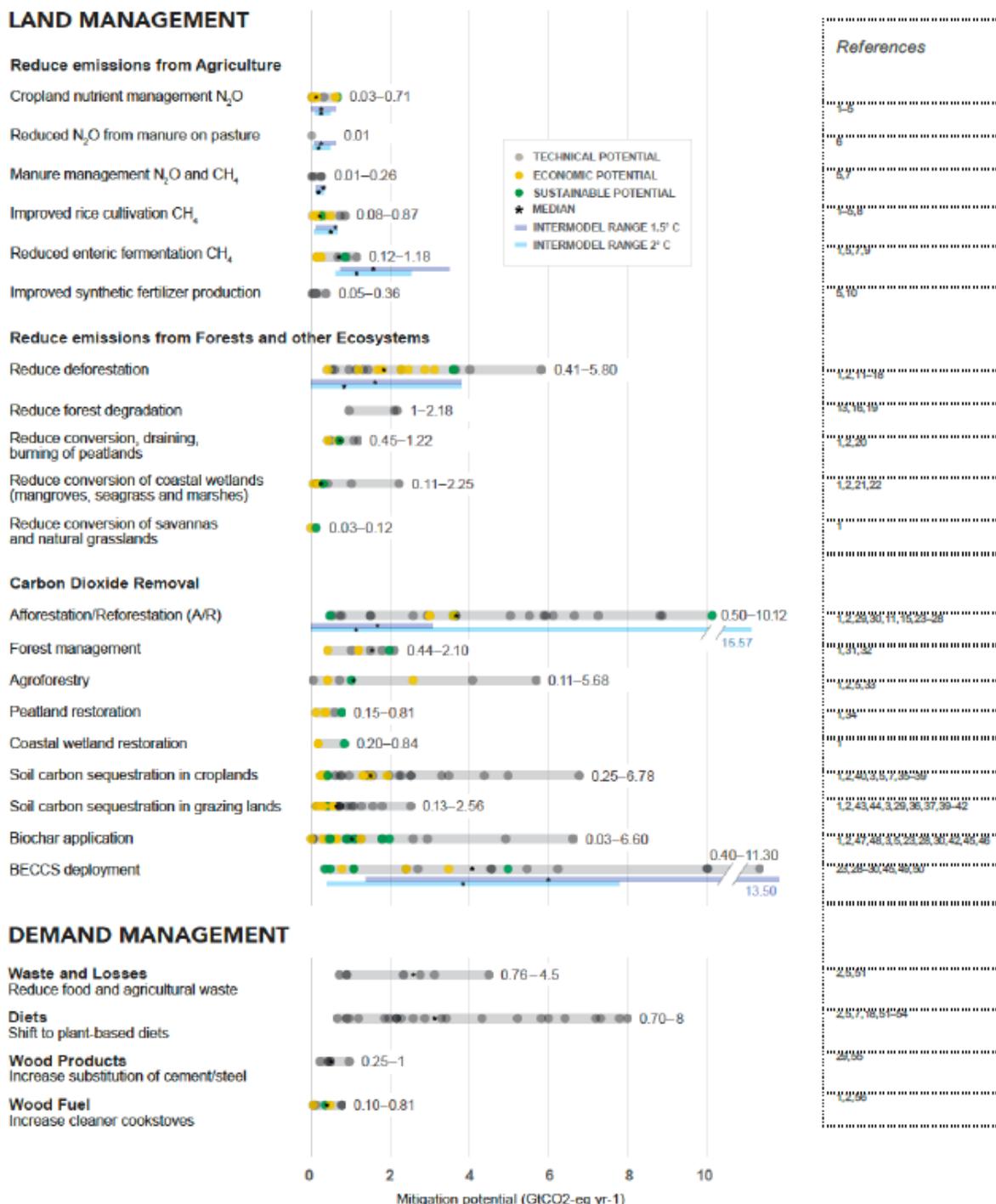
Some response options result in a net removal of GHGs from the atmosphere and storage in living or dead organic material, or in geological stores (IPCC SR15). Such options are frequently referred to in the literature as carbon dioxide removal (CDR), Greenhouse Gas Removal (GGR) or negative emissions technologies (NETs). CDR options are assessed alongside emissions reduction options. Although they have a land footprint, solar and wind farms are not assessed here as they affect greenhouse gas flux in the energy industrial sectors with minimal effect in the land sector, but the impact of solar farms on agricultural land competition is dealt with in Chapter 7.

A number of different types of scenario approach exist for estimating climate contribution of land-based response options (see Cross-Chapter Box 1: Scenarios, Chapter 1). Mitigation potentials have been estimated for single and sometimes multiple response options using stylised “bottom-up” scenarios. Response options are not mutually exclusive (e.g., management of soil carbon and cropland management). Different options interact with each other; they may have additive effects or compete with each other for land or other resources, thus these potentials cannot necessarily be added up. The interplay between different land-based mitigation options, as well as with mitigation options in other sectors (such as energy or transport), in contributing to specific mitigation pathways has been assessed using Integrated Assessment Models, see Section 2.7.2. These include interactions with wider socioeconomic conditions (see Cross-Chapter Box 1: Scenarios, Chapter 1) and other sustainability goals (see chapter 6).

2.6.1 Climate impacts of individual response options

Since AR5, there have been many new estimates of the climate impacts of single or multiple response options, summarised in Figure 2.24 and discussed in sub-sections below. Recently published syntheses of mitigation potential of land-based response options (e.g. Hawken 2017a; Smith et al. 2016b; Griscom et al. 2017a; Minx et al. 2018; Fuss et al. 2018b; Nemet et al. 2018) are also included in Figure 2.24. The wide range in mitigation estimates reflects differences in methodologies that may not be directly comparable, and estimates cannot be necessarily be added if they were calculated independently as they may be competing for land and other resources.

Some studies assess a “technical mitigation potential” - the amount possible with current technologies. Some include resource constraints (e.g., limits to yields, limits to natural forest conversion) to assess a “sustainable potential”. Some assess an “economic potential” mitigation at different carbon prices. Few include social and political constraints (e.g. behaviour change, enabling conditions, see Chapter 7), the biophysical climate effects (Section 2.5), or the impacts of future climate change (Section 2.3). Carbon stored in biomass and soils may be at risk of future climate change (see Section 2.2), natural disturbances such as wildfire (see Cross-Chapter Box 3: Fire and Climate Change, in this chapter) and future changes in land use or management changes that result in a net loss of carbon (Gren and Aklilu 2016).



1
2
3
4
5
6
7
8
9
10
11
12
13
14
15

Figure 2.24 Mitigation potential of response options in 2020-2050, measured in GtCO₂-eq yr⁻¹, adapted from Roe et al. (2017). Mitigation potentials reflect the full range of low to high estimates from studies published after 2010, differentiated according to technical (possible with current technologies), economic (possible given economic constraints) and sustainable potential (technical or economic potential constrained by sustainability considerations). Medians are calculated across all potentials in categories with >4 data points. Only includes references that explicitly provide mitigation potential estimates in CO₂-eq yr⁻¹ (or similar derivative) by 2050. Not all options land management potentials are additive as some may compete for land. Estimates reflect a range of methodologies (including definitions, global warming potentials and time horizons) that may not be directly comparable or additive. Results from Integrated Assessment Models are shown to compare with single option “bottom-up” estimates, in available categories from the 2°C and 1.5°C scenarios in the SSP Database (Version 2.0). The models reflect land management changes, yet in some instances, can also reflect demand-side effects from carbon prices, so may not be defined exclusively as “supply-side”. References: 1. Griscom et al. (2017); 2. Hawken (2017); 3. Paustian et al. (2016); 4. Beach et al. (2016); 5. Dickie et al. (2014); 6. Herrero et al. (2013); 7. Herrero et al. (2016); 8. Hussain et al. (2015); 9. Hristov, et al. (2013); 10. Zhang et al. (2013); 11. Houghton &

1 Nassikas (2018); 12. Busch & Engelmann (2017); 13. Baccini et al. (2017); 14. Zarin et al. (2016); 15. Houghton,
 2 et al. (2015); 16. Federici et al. (2015); 17. Carter et al. (2015); 18. Smith et al. (2013); 19. Pearson et al. (2017);
 3 20. Hooijer et al. (2010); 21. Howard (2017); 22. Pendleton et al. (2012); 23. Fuss et al. (2018); 24. Dooley &
 4 Kartha (2018); 25. Kreidenweis et al. (2016); 26. Yan et al. (2017); 27. Sonntag et al. (2016); 28. Lenton (2014);
 5 29. McLaren (2012); 30. Lenton (2010); 31. Sasaki et al. (2016); 32. Sasaki et al. (2012); 33. Zomer et al. (2016);
 6 34. Couwenberg et al. (2010); 35. Conant et al. (2017); 36. Sanderman et al. (2017); 37. Frank et al. (2017); 38.
 7 Henderson et al. (2015); 39. Sommer & Bossio (2014); 40. Lal (2010); 41. Zomer et al. (2017); 42. Smith et al.
 8 (2016); 43. Poeplau & Don (2015); 44. Powlson et al. (2014); 45. Powell & Lenton (2012); 46. Woolf et al. (2010);
 9 47. Roberts et al. (2010); 48. Pratt & Moran (2010); 49. Turner et al. (2018); 50. Koornneef et al. (2012); 51.
 10 Bajželj et al. (2014); 52. Springmann et al. (2016); 53. Tilman & Clark (2014); 54. Hedenus et al. (2014); 55.
 11 Miner (2010); 56. Bailis et al. (2015)

15 2.6.1.1 Land management in agriculture

16 Reducing non-CO₂ emissions from agriculture through cropland nutrient management, enteric fermentation,
 17 manure management, rice cultivation and fertiliser production has a total mitigation potential of 0.30–3.38
 18 GtCO₂-eq yr⁻¹ (*medium confidence*) (combined sub-category measures in **Figure 2.24**, details below) with a
 19 further 0.25–6.78 GtCO₂-eq yr⁻¹ from soil carbon management (Section 2.6.1.3). Other literature that looks at
 20 broader categories finds mitigation potential of 1.4–2.3 GtCO₂-eq yr⁻¹ from improved cropland management
 21 (Smith et al. 2008, 2014; Pradhan et al., 2013); 1.4–1.8 GtCO₂-eq yr⁻¹ from improved grazing land
 22 management (Conant et al. 2017; Herrero et al. 2016; Smith et al. 2008, 2014) and 0.2–2.4 GtCO₂-eq yr⁻¹
 23 from improved livestock management (Smith et al. 2008, 2014; Herrero et al. 2016, FAO 2007). A detailed
 24 discussion of the mitigation potential of agricultural response options and their co-benefits are provided in
 25 Chapter 5, Section 5.5. and 5.6).

26
 27 The three main measures to reduce enteric fermentation include improved animal diets (higher quality, more
 28 digestible livestock feed), supplements and additives (reduce methane by changing the microbiology of the
 29 rumen), and animal management and breeding (improve husbandry practices and genetics) – and applying
 30 these measures can mitigate 0.12–1.18 GtCO₂-eq yr⁻¹ (*medium confidence*) (Hristov et al. 2013; Dickie et al.
 31 2014; Herrero et al. 2016; Griscom et al. 2017). However, these measures may have limitations such as need
 32 of crop-based feed (Pradhan et al. 2013) and associated ecological costs and toxicity and animal welfare
 33 issues related to food additives (Llonch et al. 2017). Measures to manage manure include anaerobic digestion
 34 for energy use, composting as a nutrient source, reducing storage time, and changing livestock diets, and
 35 have a potential of 0.01–0.26 GtCO₂-eq yr⁻¹ (Herrero et al. 2016a; Dickie et al. 2014).

36
 37 On croplands, there is a mitigation potential of 0.03–0.71 GtCO₂-eq yr⁻¹ for cropland nutrient management
 38 (fertiliser application) (*medium confidence*) (Griscom et al. 2017a; Hawken 2017; Paustian et al. 2016;
 39 Dickie et al. 2014; Beach et al. 2015). Reducing emissions from rice production through improved water
 40 management (periodic draining of flooded fields to reduce methane emissions from anaerobic
 41 decomposition), and straw residue management (apply in dry conditions instead of on flooded fields, avoid
 42 burning to reduce methane and nitrous oxide emissions) has the potential to mitigate up to 60% of emissions
 43 (Hussain et al. 2015), or 0.08–0.87 GtCO₂-eq yr⁻¹ (*medium confidence*) (Griscom et al. 2017a; Hawken
 44 2017; Paustian et al. 2016; Hussain et al. 2015; Dickie et al. 2014; Beach et al. 2015). Further, sustainable
 45 intensification through the integration of crop and livestock systems can increase productivity, decrease
 46 emission intensity and act as a climate adaptation option (see chapter 5.5.1.4).

47
 48 Agroforestry is a land management system that combines woody biomass (e.g., trees or shrubs) with crops
 49 and/or livestock). The mitigation potential from agroforestry ranges between 0.08 to 5.7 GtCO₂ yr⁻¹, (*medium*
 50 *confidence*) (Griscom et al. 2017c; Dickie et al. 2014; Zomer et al. 2016; Hawken 2017). The high estimate
 51 is from an optimum scenario combining four agroforestry solutions (silvopasture, tree intercropping,
 52 multistrata agroforestry and tropical staple trees) of Hawken (2017a). Zomer et al. (2016) reported that the
 53 trees in agroforestry landscapes had increased carbon stock by 7.33 GtCO₂ between 2000 and 2010, or 0.7
 54 GtCO₂ yr⁻¹. For more details see Chapter 5, Section 5.5.1.3.

56 2.6.1.2 Land management in forests

57 The mitigation potential for reducing and/or halting deforestation and degradation ranges from 0.4 to 5.8

1 GtCO₂ yr⁻¹ (*high confidence*) (Griscom et al. 2017a; Hawken 2017; Busch and Engelmann 2017; Baccini et
2 al. 2017; Zarin et al. 2016b; Federici et al. 2015; Carter et al. 2015; Houghton et al. 2015; Smith et al. 2013a;
3 Houghton and Nassikas 2018). The higher figure represents a complete halting of land use conversion in
4 forests and peatlands (i.e. assuming recent rates of carbon loss are saved each year). Separate estimates of
5 degradation only range from 1.0-2.18 GtCO₂ yr⁻¹. Reduced deforestation and forest degradation include
6 conservation of existing carbon pools in vegetation and soil through protection in reserves, controlling
7 disturbances such as fire and pest outbreaks, and changing management practices. Differences in estimates
8 stem from varying land cover definitions, time periods assessed, and carbon pools included (most higher
9 estimates include belowground, dead wood, litter, soil, and peat carbon). When deforestation and
10 degradation are halted, it may take many decades to fully recover the biomass initially present in native
11 ecosystems (Meli et al. 2017, See also Chapter 4.8.3).

12
13 Afforestation/Reforestation (A/R) and forest restoration can increase carbon sequestration in both vegetation
14 and soils by 0.5–10.1 GtCO₂ yr⁻¹ (*medium confidence*) (Fuss et al. 2018; Griscom et al. 2017a; Hawken
15 2017; Kreidenweis et al. 2016; Li et al. 2016; Huang et al. 2017; Sonntag et al. 2016; Lenton 2014; McLaren
16 2012; Lenton 2010; Erb et al. 2018a; Dooley and Kartha 2018; Yan et al. 2017; Houghton et al. 2015;
17 Houghton and Nassikas 2018). Afforestation is the conversion to forest of land that historically has not
18 contained forests. Reforestation is the conversion to forest of land that has previously contained forests but
19 that has been converted to some other use. Forest restoration refers to practices aimed at regaining ecological
20 integrity in a deforested or degraded forest landscape. The lower estimate represents the lowest range from
21 an earth system model (Yan et al. 2017) and of sustainable global negative emissions potential (Fuss et al.
22 2018), and the higher estimate reforests all areas where forests are the native cover type, constrained by food
23 security and biodiversity considerations (Griscom et al. 2017a). It takes time for full carbon removal to be
24 achieved as the forest grows. Removal occurs at faster rates in young to medium aged forests and declines
25 thereafter such that older forest stands have smaller carbon removals but larger stocks with net uptake of
26 carbon slowing as forests reach maturity (Yao et al. 2018; Poorter et al. 2016; Tang et al. 2014). The land
27 intensity of afforestation and reforestation has been estimated at 0.0029 km² tC⁻¹ yr⁻¹ (Smith et al. 2016a).
28 Boysen et al. (2017) estimated that to sequester about 100 GtC by 2100 would require 13 Mkm² of
29 abandoned cropland and pastures. See also Chapter 4.8.3.

30
31 Forest management has the potential to mitigate 0.4-2.1 GtCO₂-eq yr⁻¹ (*medium confidence*) (Sasaki et al.
32 2016; Griscom et al. 2017; Sasaki et al. 2012). Forest management can alter productivity, turnover rates,
33 harvest rates carbon in soil, and carbon in wood products (Erb et al. 2017; Campioli et al. 2015; Birdsey and
34 Pan 2015; Erb et al. 2016; Noormets et al. 2015; Wäldchen et al. 2013; Malhi et al. 2015; Quesada et al.
35 2018; Nabuurs et al. 2017; Bosello et al. 2009) (see also Chapter 4, Section 4.8.4). Fertilisation may enhance
36 productivity but would increase N₂O emissions. Preserving and enhancing carbon stocks in forests has
37 immediate climate benefits but the sink can saturate and is vulnerable to future climate change (Seidl et al.
38 2017). Wood can be harvested and used for bioenergy substituting for fossil fuels (with or without carbon
39 capture and storage) (Section 2.6.1.5), for long-lived products such as timber (see below), to be buried as
40 biochar (Section 2.6.1.1) or to be used in the wider bioeconomy, enabling areas of land to be used
41 continuously for mitigation. This leads to initial carbon loss and lower carbon stocks but with each harvest
42 cycle, the carbon loss (debt) can be paid back and after a parity time, result in net savings (Laganière et al.
43 2017; Bernier and Paré 2013; Mitchell et al. 2012; Haberl et al. 2012; Haberl 2013; Ter-Mikaelian et al.
44 2015; Macintosh et al. 2015). The trade-off between maximising forest C stocks and maximising substitution
45 is highly dependent on the counterfactual assumption (no-use vs. extrapolation of current management),
46 initial forest conditions and site-specific contexts such as regrowth rates and, the displacement factors and
47 efficiency of substitution, and relative differences in emissions released during extraction, transport and
48 processing of the biomass- or fossil-based resources as well as assumptions about emission associated with
49 the product or energy source that is substituted (Grassi et al. 2018b; Nabuurs et al. 2017; Pingoud et al. 2018;
50 Smyth et al. 2017a; Luysaert et al. 2018; Valade et al. 2017; York 2012; Ter-Mikaelian et al. 2014; Naudts
51 et al. 2016b; Mitchell et al. 2012; Haberl et al. 2012; Macintosh et al. 2015; Laganière et al. 2017; Haberl
52 2013). This leads to uncertainty about optimum mitigation strategies in managed forests, while high carbon
53 ecosystems such as primary forests would have large initial carbon losses and long pay-back times and thus
54 protection of stocks would be more optimal (Lemprière et al. 2013; Kurz et al. 2016; Keith et al. 2014). See
55 also 4.8.4.

1 Global mitigation potential from increasing the demand of wood products to replace construction materials
2 range from 0.25–1 GtCO₂-eq yr⁻¹ (McLaren 2012; Miner 2010) (*medium confidence*), the uncertainty is
3 determined in part by consideration of the factors described above, and is sensitive to the displacement
4 factor, or the substitution benefit in CO₂, when wood is used instead of another material, which may vary in
5 the future as other sectors reduce emissions, as well as market factors (Sathre and O'Connor 2010; Nabuurs
6 et al. 2018; Iordan et al. 2018; Braun et al. 2016; Gustavsson et al. 2017; Peñaloza et al. 2018; Soimakallio
7 et al. 2016; Grassi et al. 2018b). Using harvested carbon in long-lived products (e.g., for construction) can
8 represent a store that can sometimes be from decades to over a century while the wood can also substitute for
9 intensive building materials, avoiding emissions from the production of concrete and steel (Sathre and
10 O'Connor 2010; Smyth et al. 2017b; Nabuurs et al. 2007; Lemprière et al. 2013). The harvest of carbon and
11 storage in products affects the net carbon balance of the forest sector, with the aim of sustainable forest
12 management strategies being to optimise carbon stocks and use of harvested products to generate sustained
13 mitigation benefits (Nabuurs et al. 2007).

14
15 Biophysical effects of forest response options are variable depending on the location and scale of activity
16 (Section 2.6). Reduced deforestation or afforestation in the tropics contributes to climate mitigation through
17 both biogeochemical and biophysical effects. It also maintains rainfall recycling to some extent. In contrast,
18 in higher latitude boreal areas observational and modelling studies show that afforestation and reforestation
19 lead to local and global warming effects, particularly in snow covered regions in the winter as the albedo is
20 lower for forests than bare snow (Bathiany et al. 2010a; Dass et al. 2013; Devaraju et al. 2018b; Ganopolski
21 et al. 2001b; Snyder et al. 2004a; West et al. 2011; Arora and Montenegro 2011) (Section 2.6). Management,
22 e.g. thinning practices in forestry, could increase the albedo in regions where albedo decreases with age. The
23 length of rotation cycles in forestry affects tree height and thus roughness, and through the removal of leaf
24 mass, harvest reduces evapotranspiration (Erb et al. 2017) which could lead to increased fire susceptibility in
25 the tropics. In temperate and boreal sites, biophysical forest management effects on surface temperature were
26 shown to be of similar magnitude than changes in land cover (Luyssaert et al. 2014b). These biophysical
27 effects could be of a magnitude to overcompensate biogeochemical effects, e.g. the sink strength of
28 regrowing forests after past depletions (Luyssaert et al. 2018; Naudts et al. 2016b), but many parameters and
29 assumptions on counterfactual influence the account (Anderson et al. 2011; Li et al. 2015b; Bright et al.
30 2015).

31
32 Forest cover also affects climate through reactive gases and aerosols with *limited evidence* and *medium*
33 *agreement* that the decrease in the emissions of BVOC resulting from the historical conversion of forests to
34 cropland has, resulted in a positive radiative forcing through direct and indirect aerosol effects, a negative
35 radiative forcing through the reduction in the atmospheric lifetime of methane it has increased an decreased
36 ozone concentrations in different regions (see Section 2.4).

37 38 39 **2.6.1.3 Land management of soils**

40 The global mitigation potential for increasing soil organic matter stocks in mineral soils is estimated to be in
41 the range of 0.4–8.64 GtCO₂ yr⁻¹ (*high confidence*) though the full literature range is wider with high
42 uncertainty related to some practices (Fuss et al. 2018; Sommer and Bossio 2014; Lal 2010; Lal et al. 2004;
43 Conant et al. 2017; Dickie et al. 2014; Frank et al. 2017a; Griscom et al. 2017c; Herrero et al. 2015, 2016b;
44 McLaren 2012; Paustian et al. 2016; Poeplau and Don 2015; Powlson et al. 2014b; Smith et al. 2016d;
45 Zomer et al. 2017). Some studies have separate potentials for soil carbon sequestration in croplands (0.25-
46 6.78 GtCO₂ yr⁻¹) (Griscom et al. 2017a; Hawken 2017; Frank et al. 2017a; Paustian et al. 2016; Herrero et al.
47 2016a; Henderson et al. 2015b; Dickie et al. 2014; Conant et al. 2017; Lal 2010) and soil carbon
48 sequestration in grazing lands (0.13-2.56 GtCO₂ yr⁻¹) (Griscom et al. 2017a; Hawken 2017; Frank et al.
49 2017a; Paustian et al. 2016; Powlson et al. 2014a; McLaren 2012; Zomer et al. 2017; Smith et al. 2015;
50 Sommer and Bossio 2014; Lal 2010). The potential for soil carbon sequestration and storage varies
51 considerably, depending on prior and current land management approaches, soil type, resource availability,
52 environmental conditions, microbial composition and nutrient availability among others (Hassink and
53 Whitmore 1997; Smith and Dukes 2013; Palm et al. 2014; Lal 2013; Six et al. 2002; Feng et al. 2013). Soils
54 are a finite carbon sink and sequestration rates may decline to negligible levels over as little as a couple of
55 decades as soils reach carbon saturation (West et al. 2004; Smith and Dukes 2013). The sink is at risk of
56 reversibility, in particular due to increased soil respiration under higher temperatures (section 2.3)

1
2 Land management practices to increase carbon interact with agricultural and fire management practices (see
3 Cross-chapter box 3: Fire and Climate Change, and Chapter 5) and include improved rotations with deeper
4 rooting cultivars, addition of organic materials, and agroforestry (Lal 2011; Smith et al. 2008; Lorenz and
5 Pitman 2014; Lal 2013; Vermeulen et al. 2012a; de Rouw et al. 2010). Adoption of green manure cover
6 crops, while increasing cropping frequency or diversity, helps sequester SOC (Poepflau and Don 2015;
7 Mazzoncini et al. 2011; Luo et al. 2010). Studies of the long-term SOC sequestration potential of
8 conservation agriculture, i.e. the simultaneous adoption of minimum tillage, (cover) crop residue retention
9 and associated soil surface coverage, and crop rotations, includes results that are both positive (Powlson et
10 al. 2016; Zhang et al. 2014) and inconclusive (Cheesman et al. 2016; Palm et al. 2014; Govaerts et al. 2009).

11
12 The efficacy of reduced and zero-till practices is highly context-specific; many studies demonstrate increased
13 carbon storage (e.g. Paustian et al., 2000; Six et al., 2004; van Kessel et al., 2013), while others show the
14 opposite effect (Sisti et al. 2004; Álvaro-Fuentes et al. 2008; Christopher et al. 2009). On the other hand,
15 deep ploughing can contribute to SOC sequestration by burying soil organic matter in the subsoil where it
16 decomposes slowly (Alcántara et al. 2016). Meta-analyses (Haddaway et al. 2017; Luo et al. 2010; Meurer
17 et al. 2018) also show a mix of positive and negative responses, and the lack of robust comparisons of soils
18 on an equivalent mass basis continues to be a problem for credible estimates (Wendt and Hauser 2013;
19 Powlson et al. 2011; Powlson et al. 2014).

20
21 Soil carbon management interacts with N₂O (Paustian et al. 2016). For example, (Li et al. 2005) estimate that
22 the management strategies required to increase C sequestration (reduced tillage, crop residue, and manure
23 recycling) would increase N₂O emissions significantly, offsetting 75–310% of the C sequestered in terms of
24 CO₂ equivalence, while other practices such as cover crops can reduce N₂O emissions (Kaye and Quemada
25 2017).

26
27 The management of soil erosion could avoid a net emissions of 1.36– 3.67 GtCO₂ yr⁻¹ and create a sink of
28 0.44–3.67 GtCO₂ yr⁻¹ (*low confidence*) (Jacinthe and Lal 2001; Lal et al. 2004; Stallard 1998; Smith et al.
29 2001; Van Oost et al. 2007). The overall impact of erosion control on mitigation is context-specific and
30 uncertain at the global level, uncertain and the final fate of eroded material is still debated (Hoffmann et al.,
31 2013).

32
33 Biochar is produced by thermal decomposition of biomass in the absence of oxygen (pyrolysis) into a stable,
34 long-lived product like charcoal that is relatively resistant to decomposition (Lehmann et al. 2015) and which
35 can stabilise organic matter added to soil (Han Weng et al. 2017). Although charcoal has been used
36 traditionally by many cultures as a soil amendment, “modern biochar”, produced in facilities that control
37 emissions, is not widely used. The range of global potential of biochar is 0.03–6.6 GtCO₂-eq yr⁻¹ by 2050
38 including energy substitution, with 0.03–4.9 GtCO₂ yr⁻¹ for CDR only (*medium confidence*) (Griscom et al.
39 2017a; Hawken 2017; Paustian et al. 2016; Fuss et al. 2018; Lenton 2014, 2010; Powell and Lenton 2012a;
40 Woolf et al. 2010; Pratt and Moran 2010; Smith 2016; Roberts et al. 2010). An analysis in which biomass
41 supply constraints were applied to protect against food insecurity, loss of habitat and land degradation,
42 estimated *technical potential* abatement of 3.7–6.6 GtCO₂-eq yr⁻¹ (including 2.6–4.6 GtCO₂ yr⁻¹ carbon
43 stabilisation) (Woolf et al. 2010). Fuss et al. (2018) propose a range of 0.5–2 GtCO₂-eq yr⁻¹ as
44 the *sustainable potential* for negative emissions through biochar. (Griscom et al. 2017b) suggest a potential
45 of 1.0 GtCO₂ yr⁻¹ based on available residues. Biochar can provide additional climate change mitigation
46 benefits by decreasing nitrous oxide (N₂O) emissions from soil and reducing nitrogen fertiliser requirements
47 in agricultural soils (Borchard et al. 2019). Application of biochar to cultivated soils can darken the surface
48 and reduce its mitigation potential via decreases in surface albedo, but the magnitude of this effect depends
49 on soil moisture content, biochar application method and type of land use (*low confidence*) (Verheijen et al.
50 2013; Bozzi et al. 2015). Biochar is discussed in more detail in Chapter 4, Section 4.9.5.

51 52 **2.6.1.4 Land management in other ecosystems**

53 Protection and restoration of wetlands, peatlands and coastal habitats reduces net carbon loss (primarily from
54 sediment/soils) and provides continued or enhanced natural CO₂ removal (Chapter 4, section 4.9.4).
55 Reducing annual emissions from peatland conversion, draining and burning could mitigate 0.45–1.22 GtCO₂-
56 eq yr⁻¹ up to 2050 (*medium confidence*) (Hooijer et al. 2010; Griscom et al. 2017; Hawken 2017) and

1 peatland restoration 0.15 to 0.81 (*low confidence*) (Couwenberg et al. 2010; Griscom et al. 2017b). The
2 upper end from Griscom et al. (2017) represents a maximum sustainable potential (accounting for
3 biodiversity and food security safeguards) for rewetting and biomass enhancement. Wetland drainage and
4 rewetting was included as a flux category under the second commitment Period of the Kyoto protocol, with
5 significant management knowledge gained over the last decade (IPCC 2013c). However, there are high
6 uncertainties as to the carbon storage and flux rates, in particular the balance between CH₄ sources and CO₂
7 sinks (Spencer et al. 2016). Peatlands are sensitive to climate change which may increase carbon uptake by
8 vegetation and carbon emissions due to respiration, with the balance being regionally dependent (*high*
9 *confidence*). There is *low confidence* about the future peatland sink globally. Some peatlands have been
10 found to be resilient to climate change (Minayeva and Sirin 2012), but the combination of land use change
11 and climate change may make them vulnerable to fire (Sirin et al. 2011). While models show mixed results
12 for the future sink (Spahni et al. 2013; Chaudhary et al. 2017; Ise et al. 2008), a study that used extensive
13 historical data sets to project change under future warming scenarios found that the currently global peatland
14 sink could increase slightly until 2100 and decline thereafter (Gallego-Sala et al. 2018).

15
16 Reducing the conversion of coastal wetlands (mangroves, seagrass and marshes) could reduce emissions by
17 0.11–2.25 GtCO₂-eq yr⁻¹ by 2050 (*medium confidence*) (Pendleton et al. 2012; Griscom et al. 2017a; Howard
18 et al. 2017; Hawken 2017). Mangrove restoration can mitigate the release of 0.07 GtCO₂ yr⁻¹ through
19 rewetting (Crooks et al. 2011) and take up 0.02–0.84 GtCO₂ yr⁻¹ from biomass and soil enhancement
20 (Griscom et al. 2017b) (*medium confidence*). The ongoing benefits provided by mangroves as a natural
21 carbon sink can be nationally-important for Small Island Developing States (SIDS) and other countries with
22 extensive coastlines, based on estimates of high carbon sequestration rates per unit area (McLeod et al. 2011;
23 Duarte et al. 2013; Duarte 2017; Taillardat et al. 2018). There is only *medium confidence* in the effectiveness
24 of enhanced carbon uptake using mangroves, due to the many uncertainties regarding the response of
25 mangroves to future climate change (Jennerjahn et al. 2017); dynamic changes in distributions (Kelleway et
26 al. 2017) and other local-scale factors affecting long-term sequestration and climatic benefits (e.g., methane
27 release; Dutta et al. 2017). The climate mitigation potential of coastal vegetated habitats (mangrove forests,
28 tidal marshes and seagrasses) is considered in Chapter 5 of the IPCC Special Report on the Ocean,
29 Cryosphere and Climate Change (SROCC), in a wider ‘blue carbon’ context.

30 31 32 **2.6.1.5 Bioenergy and bioenergy with carbon capture and storage (BECCS)**

33 An introduction and overview of bioenergy and BECCS can be found in the Cross-Chapter Boxes: Cross-
34 Chapter Box 12: Traditional biomass use, Chapter 7; Cross-Chapter Box 7: Bioenergy and BECCS in
35 mitigation scenarios, Chapter 6. CCS technologies are discussed in SR15. The discussion below refers to
36 modern bioenergy only, (e.g., liquid biofuels for transport and the use of solid biofuels in combined heat and
37 power plants).

38
39 The mitigation potential of bioenergy coupled with CCS, i.e., BECCS, is estimated to be between 0.4 and
40 11.3 GtCO₂ yr⁻¹ (*medium confidence*) based on studies that directly estimate mitigation for BECCS (not
41 bioenergy) in units of CO₂ (not EJ) (McLaren 2012; Lenton 2014; Fuss et al. 2018; Turner et al. 2018b;
42 Lenton 2010; Koornneef et al. 2012; Powell and Lenton 2012b). SR15 reported a potential of 1-85 GtCO₂ yr⁻¹
43 ¹ which they noted could be narrowed to a range of 0.5 to 5 GtCO₂ yr⁻¹ when taking account of sustainability
44 aims (Fuss et al. 2018). The upper end of the SR15 range is considered as a theoretical potential. Previously,
45 the IPCC Special Report on Renewable Energy Sources concluded the technical potential of biomass supply
46 for energy (without BECCS) could reach 100-300 EJ yr⁻¹ by 2050, which would be 2-15 GtCO₂ yr⁻¹ (using
47 conversion factors 1 EJ = 0.02-0.05 GtCO₂ yr⁻¹ emission reduction, SR15). A range of recent studies
48 including sustainability or economic constraints estimate that 50-244 EJ (1 to 12 GtCO₂ yr⁻¹ using
49 conversion factors above) of bioenergy could be produced on 0.1-13 Mkm² of land (Fuss et al. 2018; Chan
50 and Wu 2015; Schueler et al. 2016; Wu et al. 2013; Searle and Malins 2015; Wu et al. 2019; Heck et al.
51 2018; Fritz et al. 2013) SR15 SPM).

52
53 There is *high confidence* that the most important factors determining future biomass supply for energy are
54 land availability and land productivity (Berndes et al. 2013; Creutzig et al. 2015a; Woods et al. 2015;
55 Daioglou et al. 2019). Estimates of marginal/degraded lands currently considered available for bioenergy
56 range from 3.2 to 14.0 Mkm², depending on the adopted sustainability criteria, land class definitions, soil

1 conditions, land mapping method and environmental and economic considerations (Campbell et al. 2008;
2 Cai et al. 2011; Lewis and Kelly 2014).

3
4 Bioenergy production systems can lead to net emissions in the short term that can be “paid-back” over time,
5 with multiple harvest cycles and fossil fuel substitution, unlike fossil carbon emissions (Campbell et al.
6 2008; Cai et al. 2011; Lewis and Kelly 2014; De Oliveira Bordonal et al. 2015). Stabilising bioenergy crops
7 in previous high carbon forestland or peatland results in high emissions of carbon that may take from
8 decades to more than a century to be re-paid in terms of net CO₂ emission savings from replacing fossil
9 fuels, depending on previous forest carbon stock, bioenergy yields, and displacement efficiency (Elshout et
10 al. 2015; Harper et al. 2018; Daioglou et al. 2017). In the case of bioenergy from managed forests, the
11 magnitude and timing of the net mitigation benefits is controversial, as it varies with differences due to local
12 climate conditions, forest management practice, fossil fuel displacement efficiency and methodological
13 approaches (Hudiburg et al. 2011; Berndes et al. 2013; Guest et al. 2013; Lamers and Junginger 2013;
14 Cherubini et al. 2016; Cintas et al. 2017; Laurance et al. 2018; Valade et al. 2018; Baker et al. 2019).
15 Suitable bioenergy crops can be integrated in agricultural landscapes to reverse ecosystem carbon depletion
16 (Creutzig et al. 2015a; Robertson et al. 2017; Vaughan et al. 2018; Daioglou et al. 2017). Cultivation of short
17 rotation woody crops and perennial grasses on degraded land or cropland previously used for annual crops
18 typically accumulate carbon in soils due to their deep root systems (Don et al. 2012; Robertson et al. 2017).
19 The use of residues and organic waste as bioenergy feedstock can mitigate land use change pressures
20 associated with bioenergy deployment, but residues are limited and the removal of residues that would
21 otherwise be left on the soil could lead soil degradation (Chum et al. 2011; Liska et al. 2014; Monforti et al.
22 2015; Zhao et al. 2015; Daioglou et al. 2016).

23
24 The steps required to cultivate, harvest, transport, process and use biomass for energy generate emissions of
25 GHGs and other climate pollutants (Chum et al. 2011; Creutzig et al. 2015b; Staples et al. 2017; Daioglou et
26 al. 2019). Life-cycle GHG emissions of modern bioenergy alternatives are usually lower than those for fossil
27 fuels (*robust evidence, medium agreement*) (Chum et al. 2011; Creutzig et al. 2015b). The magnitude of
28 these emissions largely depends on location, (e.g. soil quality, climate), prior land use, feedstock used (e.g.,
29 residues, dedicated crops, algae), land use practice (e.g., soil management, fertiliser use), biomass transport
30 (distances and transport modes), the bioenergy conversion pathway and product (e.g., wood pellets, ethanol).
31 Use of conventional food and feed crops as a feedstock generally provides the highest bioenergy yields per
32 hectare, but also causes more GHG emissions per unit energy compared to agriculture residues, biomass
33 from managed forests, and lignocellulosic crops such as short-rotation coppice and perennial grasses (Chum
34 et al. 2011; Gerbrandt et al. 2016). This is due to the application of fertilisers and other inputs (Oates et al.
35 2016; Rowe et al. 2016; Lai et al. 2017; Robertson et al. 2017).

36
37 Bioenergy from dedicated crops are in some cases held responsible for GHG emissions resulting from
38 indirect land use change (iLUC), that is the bioenergy activity may lead to displacement of agricultural or
39 forest activities into other locations, driven by market-mediated effects. Other mitigation options may also
40 cause iLUC. At a global level of analysis, indirect effects are not relevant because all land-use emissions are
41 direct. iLUC emissions are potentially more significant for crop-based feedstocks such as corn, wheat and
42 soybean, than for advanced biofuels from lignocellulosic materials (Chum et al. 2011; Wicke et al. 2012;
43 Valin et al. 2015; Ahlgren and Di Lucia 2014). Estimates of emissions from iLUC are inherently uncertain,
44 are widely debated in the scientific community, and are highly dependent on modelling assumptions, such as
45 supply/demand elasticities, productivity estimates, incorporation or exclusion of emission credits for
46 coproducts, scale of biofuel deployment (Rajagopal and Plevin 2013; Finkbeiner 2014; Kim et al. 2014;
47 Zilberman 2017). In some cases, iLUC effects are estimated to result in emission reductions. For example,
48 market-mediated effects of bioenergy in North America showed potential for increased carbon stocks by
49 inducing conversion of pasture or marginal land to forestland (Cintas et al. 2017; Duden et al. 2017; Dale et
50 al. 2017; Baker et al. 2019). There is a wide range of variability in iLUC values for different types of
51 biofuels, from -75 to +55 g CO₂ MJ⁻¹ (Ahlgren and Di Lucia 2014; Valin et al. 2015; Plevin et al. 2015;
52 Taheripour and Tyner 2013; Bento and Klotz 2014). There is low confidence in attribution of emissions from
53 iLUC to bioenergy.

54
55 Bioenergy deployment can have large biophysical effects on regional climate, with the direction and
56 magnitude of the impact depending on the type of bioenergy crop, previous land use and seasonality (*limited*

1 *evidence, medium agreement*). A study of two alternative future bioenergy scenarios using 15 Mkm² of
2 intensively used managed land or conversion of natural areas showed a nearly neutral effect on surface
3 temperature at global levels (considering biophysical effects and CO₂ and N₂O fluxes from land but not
4 substitution effects), although there were significant seasonal and regional differences (Kicklighter et al.
5 2013). Modelling studies on biofuels in the US found the switch from annual crops to perennial bioenergy
6 plantations like Miscanthus in the US could lead to regional cooling due to increases in evapotranspiration
7 and albedo (Georgescu et al. 2011; Harding et al. 2016), with perennial bioenergy crop expansion over
8 suitable abandoned and degraded farmlands causing near-surface cooling up to 5°C during the growing
9 season (Wang et al. 2017b). Similarly, growing sugarcane on existing cropland in Brazil cools down the
10 local surface during daytime conditions up to -1°C, but warmer conditions occurs if sugar cane is deployed at
11 the expense of natural vegetation (Brazilian Cerrado) (Loarie et al. 2011). In general, bioenergy crops (as for
12 all crops) induce a cooling of ambient air during the growing season, but after harvest the decrease in
13 evapotranspiration can induce warming (Harding et al. 2016; Georgescu et al. 2013; Wang et al. 2017b).
14 Bioenergy crops were found to cause increased isoprene emissions in a scenario where 0.69 Mkm² of oil
15 palm for biodiesel in the tropics and 0.92 Mkm² of short rotation coppice (SRC) in the mid-latitudes were
16 planted, but effects on global climate were negligible (Ashworth et al. 2012).
17

18 **2.6.1.6 Enhanced weathering**

19 Weathering is the natural process of rock decomposition via chemical and physical processes in which CO₂
20 is removed from the atmosphere and converted to bicarbonates and/or carbonates (IPCC 2005). Formation of
21 calcium carbonates in the soil provides a permanent sink for mineralised organic carbon (Manning 2008;
22 Beerling et al. 2018). Mineral weathering can be enhanced through grinding up rock material to increase the
23 surface area, and distributing it over land to provide carbon removals of 0.5–4.0 GtCO₂ yr⁻¹ (*medium*
24 *confidence*) (Beerling et al. 2018; Lenton 2010; Smith et al. 2016a; Taylor et al. 2016). While the
25 geochemical potential is quite large, agreement on the technical potential is low due to a variety of unknown
26 parameters and of limits such as rates of mineral extraction, grinding, delivery, and challenges with scaling
27 and deployment.
28

29 **2.6.1.7 Demand management in the food sector (diet change, waste reduction)**

30 Demand-side management has the potential for climate change mitigation via reducing emissions from
31 production, switching to consumption of less emission intensive commodities, and making land available for
32 carbon dioxide removal (see Chapter 5 Section 5.5.2). Reducing food losses and waste increases the overall
33 efficiency of food value chains (less land and inputs needed) along the entire supply chain, and has the
34 potential to mitigate 0.8-4.5 GtCO₂-eq yr⁻¹ (*high confidence*) (Chapter 5 section 5.5.2.5, Bajželj et al. 2014;
35 Dickie et al 2014; Hawken 2017; Hiç et al. 2016).
36

37 Shifting to diets that are lower in emissions-intensive foods like beef delivers a mitigation potential of 0.7-
38 8.0 GtCO₂-eq yr⁻¹ (*high confidence*) (Bajželj et al. 2014; Dickie et al. 2014; Herrero et al. 2016; Hawken
39 2017; Springmann et al. 2016; Tilman and Clark 2014; Hedenus et al. 2014; Stehfest et al. 2009) with most
40 of the higher end estimates (> 6 GtCO₂-eq yr⁻¹) based on veganism, vegetarianism or very low ruminant meat
41 consumption) (see Chapter 5, Section 5.5.2). In addition to direct mitigation gains, decreasing meat
42 consumption, primarily of ruminants, and reducing wastes further reduces water use, soil degradation,
43 pressure on forests, land used for feed potentially freeing up land for mitigation (Tilman and Clark 2014)
44 (see chapters 5 and 6). Additionally, consumption of locally produced food, shortening the supply chain, can
45 in some cases minimise food loss, contribute to food security, and reduce GHG emissions associated with
46 energy consumption and food loss (see Chapter 5 Section 5.5.2.6).
47

48 **2.6.2 Integrated pathways for climate change mitigation**

49 Land-based response options have the potential to interact, resulting in additive effects (e.g., climate co-
50 benefits) or negating each other (e.g., through competition for land), they also interact with mitigation
51 options in other sectors (such as energy or transport), thus they need to be assessed collectively under
52 different climate mitigation targets and in combination with other sustainability goals (Popp et al. 2017;
53 Obersteiner et al. 2016; Humpenöder et al. 2018). Integrated Assessment Models (IAMs) with distinctive
54 land-use modules are the basis for the assessment of mitigation pathways as they combine insights from
55 various disciplines in a single framework and cover the largest sources of anthropogenic GHG emissions
56

1 from different sectors (see also SR15 Chapter 2 and technical Annex for more details). IAMs consider a
2 limited, but expanding, portfolio of land-based mitigation options. Furthermore, the inclusion and detail of a
3 specific mitigation measure differs across IAMs and studies (see also SR15 and Chapter 6). For example, the
4 IAM scenarios based on the Shared Socio-economic Pathways (SSPs) (Riahi et al. 2017)(see more details on
5 the SSPs in Cross-Chapter Box 1: Scenarios, Chapter 1) include possible trends in agriculture and land use
6 for five different socioeconomic futures, but cover a limited set of land-based mitigation options: dietary
7 changes, higher efficiency in food processing (especially in livestock production systems), reduction of food
8 waste, increasing agricultural productivity, methane reductions in rice paddies, livestock and grazing
9 management for reduced methane emissions from enteric fermentation, manure management, improvement
10 of N-efficiency, 1st generation biofuels, reduced deforestation, afforestation, 2nd generation bioenergy crops
11 and BECCS (Popp et al. 2017). However, many “natural climate solutions” (Griscom et al. 2017b), such as
12 forest management, rangeland management, soil carbon management or wetland management, are not
13 included in most of these scenarios. In addition, most IAMs neglect the biophysical effects of land-use such
14 as changes in albedo or evapotranspiration with few exceptions (Kreidenweis et al. 2016).

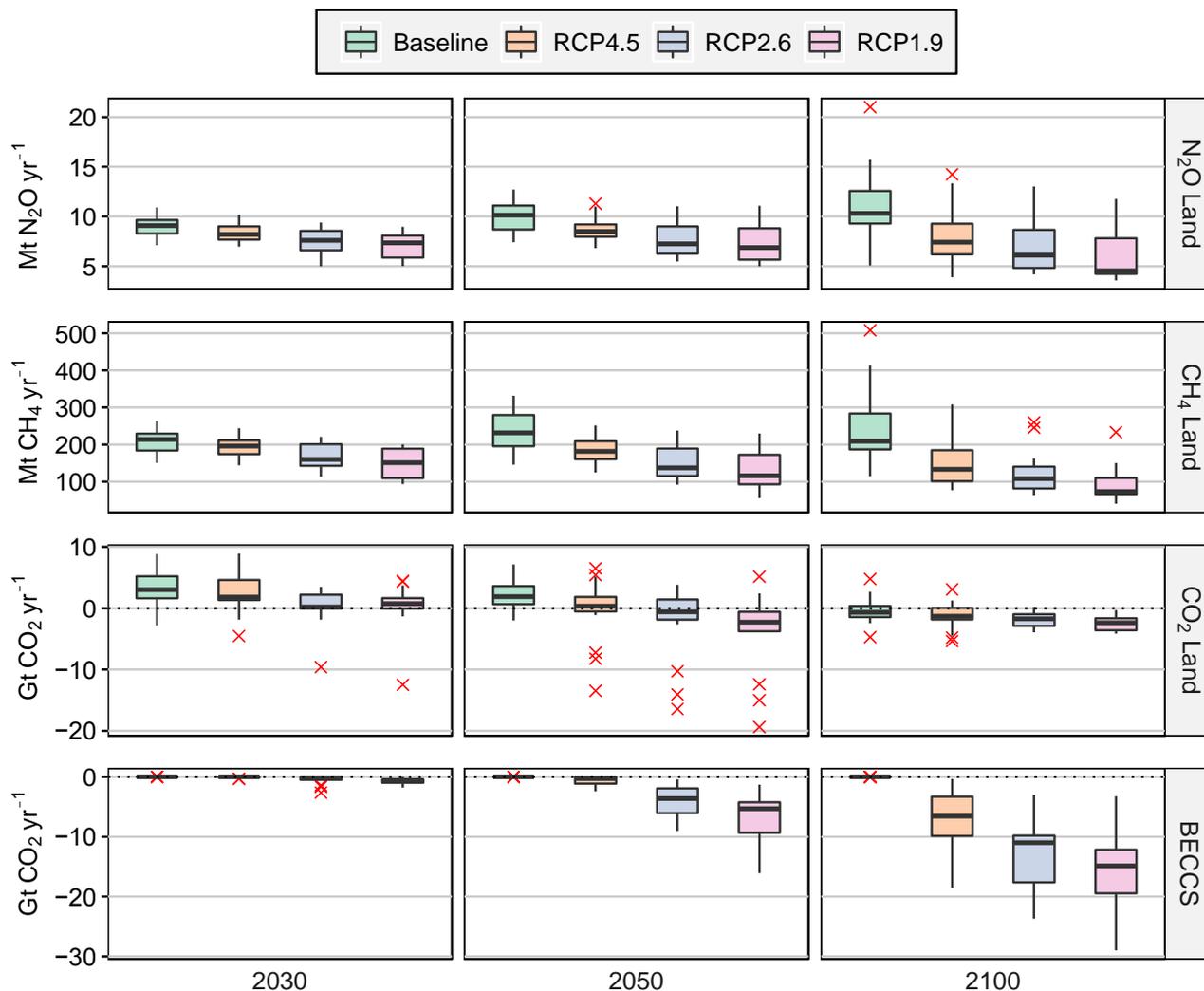
15
16 Mitigation pathways, based on IAMs, are typically designed to find the least cost pathway to achieve a pre-
17 defined climate target (Riahi et al. 2017). Such cost-optimal mitigation pathways, especially in RCP2.6
18 (broadly a 2°C target) and 1.9 scenarios (broadly a 1.5°C target), project GHG emissions to peak early in the
19 21st century, strict GHG emission reduction afterwards and, depending on the climate target, net carbon
20 dioxide removal (CDR) from the atmosphere in the second half of the century (see Chapter 2 of SR15,
21 (Tavoni et al. 2015; Riahi et al. 2017). In most of these pathways, land use is of great importance because of
22 its mitigation potential as discussed in section 2.7.1: these pathways are based on the assumptions that large-
23 scale afforestation and reforestation removes substantial amounts of CO₂ from the atmosphere; biomass
24 grown on cropland or from forestry residues can be used for energy generation or BECCS substituting fossil
25 fuel emissions and generating CDR; non-CO₂ emissions from agricultural production can be reduced, even
26 under improved agricultural management (Popp et al. 2017; Rogelj et al. 2018a; Van Vuuren et al. 2018,
27 Frank et al. 2018).

28
29 From the IAM scenarios available to this assessment, a set of feasible mitigation pathways has been
30 identified which is illustrative of the range of possible consequences on land use and GHG emissions
31 (presented in this chapter) and sustainable development (see Chapter 6). Thus, the IAM scenarios selected
32 here vary due to underlying socio-economic and policy assumptions, mitigation options considered, long-
33 term climate goal, the level of inclusion of other sustainability goals (such as land and water restrictions for
34 biodiversity conservation or food production), and models by which they are generated.

35
36 In the baseline case without climate change mitigation, global CO₂ emissions from land-use change decrease
37 over time in most scenarios due to agricultural intensification and decreases in demand for agricultural
38 commodities – some turning even negative by the end of the century due to abandonment of agricultural land
39 and associated carbon uptake through vegetation regrowth. Median global CO₂ emissions from land-use
40 change across 5 SSPs and 5 IAMs decrease throughout the 21st century: 3, 1.9 and -0.7 GtCO₂ yr⁻¹ in 2030,
41 2050 and 2100 respectively (Figure 2.25). In contrast, CH₄ and N₂O emissions from agricultural production
42 remain rather constant throughout the 21st century (CH₄: 214, 231.7 and 209.1 Mt CH₄ yr⁻¹ in 2030, 2050
43 and 2100 respectively; N₂O: 9.1, 10.1 and 10.3 Mt N₂O yr⁻¹ in 2030, 2050 and 2100 respectively).

44
45 In the mitigation cases (RCP4.5, RCP2.6 and RCP1.9), most of the scenarios indicate strong reductions in
46 CO₂ emissions due to i) reduced deforestation and ii) carbon uptake due to afforestation. However, CO₂
47 emissions from land use can occur in some mitigation scenarios as a result of weak land-use change
48 regulation (Fujimori et al. 2017; Calvin et al. 2017) or displacement effects into pasture land caused by high
49 bioenergy production combined with forest protection only (Popp et al. 2014). The level of carbon dioxide
50 removal globally (median value across SSPs and IAMs) increases with the stringency of the climate target
51 (RCP4.5, RCP2.6 and RCP1.9) for both afforestation (-1.3, -1.7 and -2.4 GtCO₂ yr⁻¹ in 2100) and BECCS (-
52 6.5, -11 and -14.9 GtCO₂ yr⁻¹ in 2100; see also Cross-Chapter Box 7: Bioenergy and BECCS in mitigation
53 scenarios, Chapter 6). In the mitigation cases (RCP4.5, RCP2.6 and RCP1.9), CH₄ and N₂O emissions are
54 remarkably lower compared to the baseline case (CH₄: 133.2, 108.4 and 73.5 Mt CH₄ yr⁻¹ in 2100; N₂O: 7.4,
55 6.1 and 4.5 Mt N₂O yr⁻¹ in 2100; see previous paragraph for CH₄ and N₂O emissions in the baseline case).
56 The reductions in the mitigation cases are mainly due to improved agricultural management such as

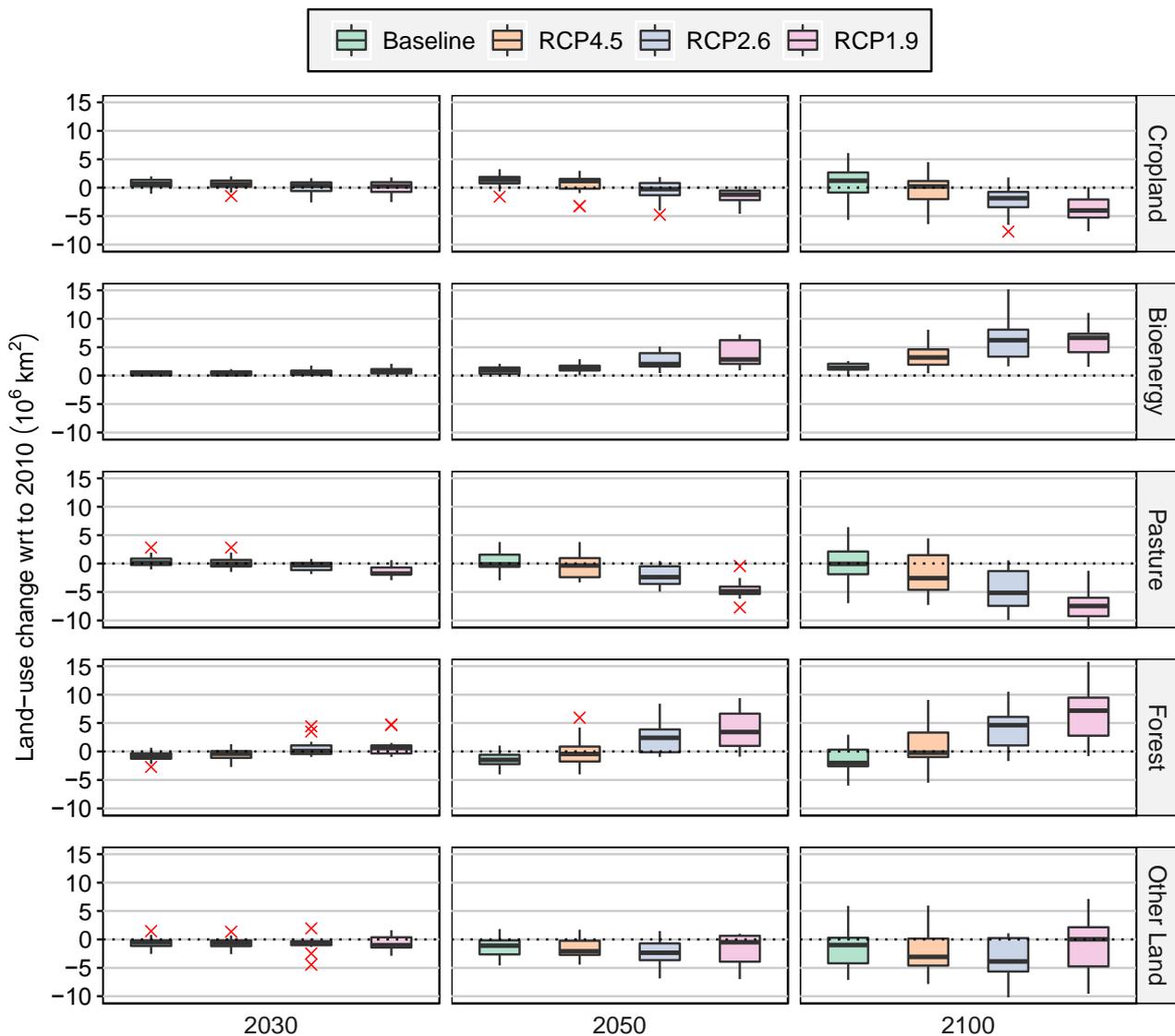
1 improved nitrogen fertiliser management, improved water management in rice production, improved manure
 2 management by for example covering of storages or adoption of biogas plants, better herd management and
 3 better quality of livestock through breeding and improved feeding practices. In addition, dietary shifts away
 4 from emission-intensive livestock products also lead to decreased CH₄ and N₂O emissions especially in
 5 RCP2.6 and RCP1.9 scenarios. However, high levels of bioenergy production can result in increased N₂O
 6 emissions due to nitrogen fertilisation of dedicated bioenergy crops.
 7



8
 9
 10 **Figure 2.25 Land-based global GHG emissions and removals in 2030, 2050 and 2100 for Baseline, RCP4.5,**
 11 **RCP2.6 and RCP1.9 based on the Shared Socioeconomic Pathways (SSP) (Popp et al. 2017; Rogelj et al.**
 12 **2018; Riahi et al. 2017). Data is from an update of the IAMC Scenario Explorer developed for the SR15**
 13 **(Huppmann et al. 2018; Rogelj et al. 2018). Boxplots (Tukey style) show median (horizontal line),**
 14 **interquartile range (IQR box) and the range of values within 1.5 x IQR at either end of the box (vertical**
 15 **lines) across 5 SSPs and across 5 IAMs. Outliers (red crosses) are values greater than 1.5 x IQR at either**
 16 **end of the box. The categories CO₂ Land, CH₄ Land and N₂O Land include GHG emissions from land-use**
 17 **change and agricultural land use (including emissions related to bioenergy production). In addition, the**
 18 **category CO₂ Land includes negative emissions due to afforestation. BECCS reflects the CO₂ emissions**
 19 **captured from bioenergy use and stored in geological deposits.**
 20
 21
 22

23 Such high levels of carbon dioxide removal through mitigation options that require land conversion (BECCS
 24 and afforestation) shape the land system dramatically (Figure 2.26). Across the different RCPs, SSPs and
 25 IAMs median change of global forest area throughout the 21st century ranges from about -0.2 to +7.2 Mkm²
 26 between 2010 and 2100, and agricultural land used for 2nd generation bioenergy crop production ranges from

1 about 3.2–6.6 Mkm² in 2100 (Popp et al. 2017; Rogelj et al. 2018). Land requirements for bioenergy and
 2 afforestation for a RCP1.9 scenario are higher than for a RCP2.6 and especially a RCP4.5 mitigation
 3 scenario. As a consequence of the expansion of mainly land-demanding mitigation options, global pasture
 4 land is reduced in most mitigation scenarios much stronger compared to baseline scenarios (median
 5 reduction of 0, 2.6, 5.1 and 7.5 Mkm² between 2010 and 2100 in Baseline, RCP4.5, RCP2.6 and RCP1.9
 6 respectively). In addition, cropland for food and feed production decreases with the stringency of the climate
 7 target (+1.2, +0.2, -1.8 and -4 Mkm² in 2100 compared to 2010 in Baseline, RCP4.5, RCP2.6 and RCP1.9
 8 respectively). These reductions in agricultural land for food and feed production are facilitated by
 9 agricultural intensification on agricultural land and in livestock production systems (Popp et al. 2017) but
 10 also by changes in consumption patterns (Fujimori et al. 2017; Frank et al. 2017b). The pace of projected
 11 land-use change over the coming decades in ambitious mitigation scenarios goes well beyond historical
 12 changes in some instances (Turner et al. 2018c), see also SR15). This raises issues for societal acceptance,
 13 and distinct policy and governance for avoiding negative consequences for other sustainability goals
 14 (Humpenöder et al. 2018; Obersteiner et al. 2016; Calvin et al. 2014), see Chapter 6 and 7).
 15
 16



17
 18 **Figure 2.26** Global change of major land cover types by 2030, 2050 and 2100 relative to 2010 for Baseline,
 19 RCP4.5, RCP2.6 and RCP1.9 based on the Shared Socioeconomic Pathways (SSP) (Popp et al. 2017; Rogelj et
 20 al. 2018; Riahi et al. 2017). Data is from an update of the IAMC Scenario Explorer developed for the SR15
 21 (Huppmann et al. 2018; Rogelj et al. 2018). Boxplots (Tukey style) show median (horizontal line),
 22 interquartile range IQR (box) and the range of values within 1.5 x IQR at either end of the box (vertical lines)
 23 across 5 SSPs and across 5 IAMs. Outliers (red crosses) are values greater than 1.5 x IQR at either end of the
 24 box. In 2010, total land cover at global scale was estimated 15-16 Mkm² for cropland, 0-0.14 Mkm² for

1 **bioenergy, 30-35 Mkm² for pasture and 37-42 Mkm² for forest, across the IAMs that reported SSP pathways**
2 **(Popp et al. 2017).**
3

4 Different mitigation strategies can achieve the net emissions reductions that would be required to follow a
5 Pathway that limits global warming to 2°C or 1.5°C, with very different consequences on the land system.
6

7 **Figure 2.27** shows six alternative pathways (archetypes) for achieving ambitious climate targets (RCP2.6
8 and RCP1.9) highlighting land-based strategies and GHG emission. All pathways are assessed by different
9 models but are all based on the Shared Socioeconomic Pathway 2 (SSP2) (Riahi et al. 2017), with all based
10 on an RCP 1.9 mitigation pathway expect for Pathway 1, which is RCP2.6. All scenarios show land-based
11 negative emissions but the amount varies across pathways, as do the relative contributions of different land-
12 based Carbon Dioxide Removal (CDR) options, such as afforestation/reforestation and bioenergy with
13 carbon capture and storage (BECCS).
14

15 Pathway 1 RCP2.6 “Portfolio” (Fricko et al. 2017) shows a strong near-term decrease of CO₂ emissions from
16 land-use change, mainly due to reduced deforestation, as well as slightly decreasing N₂O and CH₄ emissions
17 after 2050 from agricultural production due to improved agricultural management and dietary shifts away
18 from emissions-intensive livestock products. However, in contrast to CO₂ emissions, which turn net-negative
19 around 2050 due to afforestation/reforestation, CH₄ and N₂O emissions persist throughout the century due to
20 difficulties of eliminating these residual emissions based on existing agricultural management methods
21 (Stevanović et al. 2017; Frank et al. 2017b). In addition to abating land-related GHG emissions as well as
22 increasing the terrestrial sink, this example also shows the importance of the land sector in providing
23 biomass for BECCS and hence CDR in the energy sector. In this scenario, annual BECCS-based CDR is
24 about 3-times higher than afforestation-based CDR in 2100 (-11.4 and -3.8 GtCO₂ yr⁻¹ respectively).
25 Cumulative CDR throughout the century amounts to -395 GtCO₂ for BECCS and -73 GtCO₂ for
26 afforestation. Based on these GHG dynamics, the land sector turns GHG emission neutral in 2100. However,
27 accounting also for BECCS-based CDR taking place in the energy sector but with biomass provided by the
28 land sector turns the land sector GHG emission neutral already in 2060, and significantly net-negative by the
29 end of the century.
30

31 Pathway 2 RCP1.9 “Increased Ambition” (Rogelj et al. 2018) has dynamics of land-based GHG emissions
32 and removals that are very similar to those in Pathway 1 (RCP2.6) but all GHG emission reductions as well
33 as afforestation/reforestation and BECCS-based CDR start earlier in time at a higher rate of deployment.
34 Cumulative CDR throughout the century amounts to -466 GtCO₂ for BECCS and -117 GtCO₂ for
35 afforestation.
36

37 Pathway 3 RCP 1.9 “Only BECCS”, in contrast to Pathway 2, includes only BECCS-based CDR (Kriegler
38 et al. 2017). In consequence, CO₂ emissions are persistent much longer, predominantly from indirect land-
39 use change due to large-scale bioenergy cropland expansion into non-protected natural areas (Popp et al.
40 2017; Calvin et al. 2014). While annual BECCS CDR rates in 2100 are similar to Pathway 1 and 2 (-15.9
41 GtCO₂ yr⁻¹), cumulative BECCS-based CDR throughout the century is much larger (-944 GtCO₂).
42

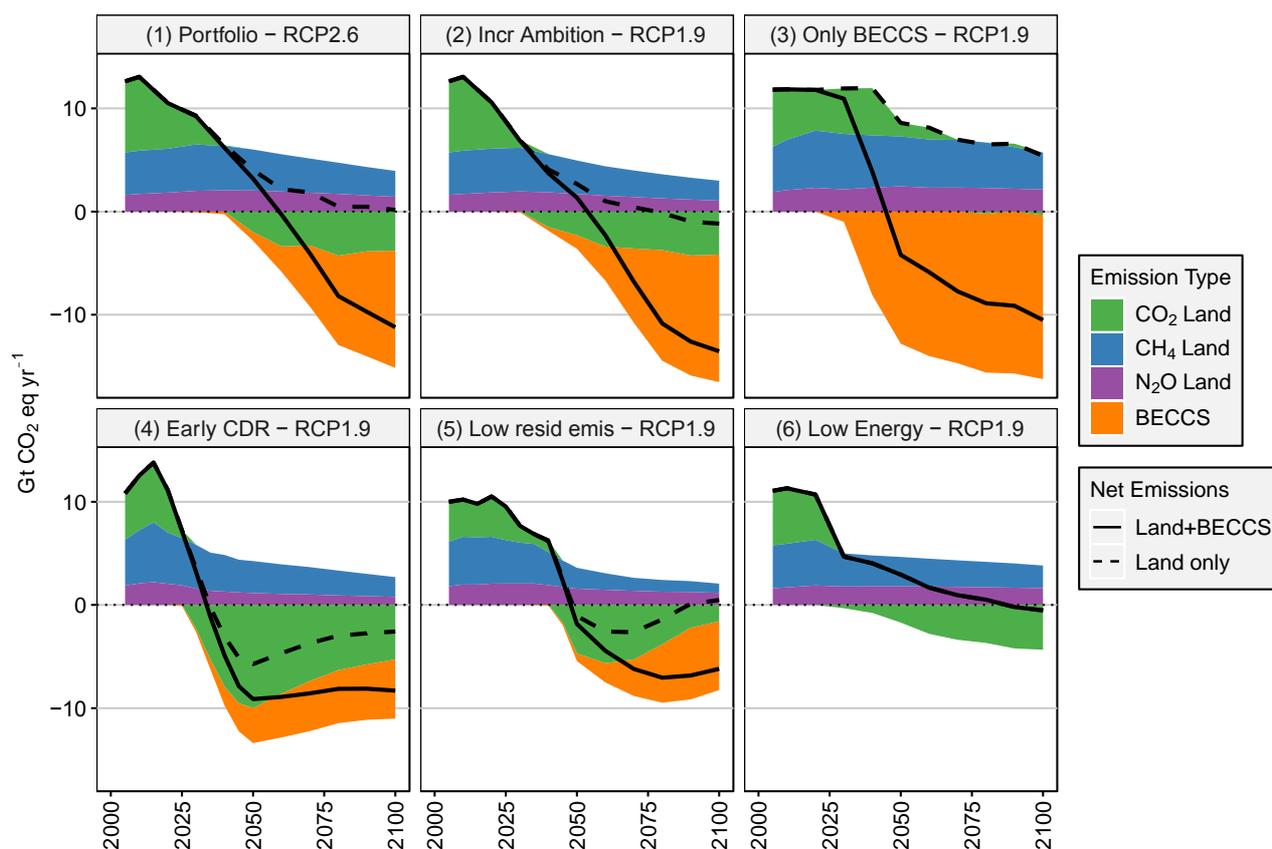
43 Pathway 4 RCP1.9 “Early CDR” (Bertram et al. 2018) indicates that a significant reduction in the later
44 century in the BECCS-related CDR as well as CDR in general can be achieved with earlier and mainly
45 terrestrial CDR, starting already in 2030. In this scenario, terrestrial CDR is based on afforestation but could
46 also be supported by soil organic carbon sequestration (Paustian et al. 2016) or other natural climate
47 solutions such as rangeland or forest management (Griscom et al. 2017b). This scenario highlights the
48 importance of the timing for CDR-based mitigation pathways (Obersteiner et al. 2016). As a result of near-
49 term and mainly terrestrial CDR deployment, cumulative BECCS-based CDR throughout the century is
50 limited to -300 GtCO₂, while cumulative afforestation-based CDR amounts to -428 GtCO₂.
51

52 In Pathway 5 RCP1.9 “Low residual emissions” (van Vuuren et al. 2018), land-based mitigation is driven by
53 stringent enforcement of measures and technologies to reduce end-of-pipe non-CO₂ emissions and by
54 introduction of in-vitro (cultured) meat, reducing residual N₂O and CH₄ emissions from agricultural
55 production. In consequence, much lower amounts of CDR from afforestation and BECCS are needed with
56 much later entry points to compensate for residual emissions. Cumulative CDR throughout the century

1 amounts to -252 GtCO₂ for BECCS and -128 GtCO₂ for afforestation. Therefore, total cumulative land-based
 2 CDR in Pathway 5 is substantially lower compared to Pathways 2-4 (380 GtCO₂).
 3

4 Finally, Pathway 6 RCP1.9 “Low Energy” (Grubler et al. 2018) – equivalent to pathway LED in SR15 –
 5 indicates the importance of other sectoral GHG emission reductions for the land sector. In this example,
 6 rapid and early reductions in energy demand and associated drops in energy-related CO₂ emissions, limit
 7 overshoot and decrease the requirements for negative emissions technologies, especially for land-demanding
 8 CDR such as biomass production for BECCS and afforestation. While BECCS is not used at all in Pathway
 9 6, cumulative CDR throughout the century for afforestation amounts to -124 GtCO₂.
 10

11 Besides their consequences on mitigation pathways and land consequences, those archetypes can also affect
 12 multiple other sustainable development goals that provide both challenges and opportunities for climate
 13 action (see Chapter 6).
 14
 15



16
 17
 18 **Figure 2.27 Evolution and break down of global land-based GHG emissions and removals under six**
 19 **alternative mitigation pathways, which illustrate the differences in timing and magnitude of land-based**
 20 **mitigation approaches including afforestation and BECCS. All pathways are based on different IAM**
 21 **realisations of SSP2. Pathway 1 is based on RCP 2.6, while all other pathways are based on RCP 1.9.**
 22 **Pathway 1: MESSAGE-GLOBIOM (Fricko et al. 2017); Pathway 2: MESSAGE-GLOBIOM (Rogelj et**
 23 **al. 2018); Pathway 3: REMIND-MAGPIE (Kriegler et al. 2017); Pathway 4: REMIND-MAGPIE**
 24 **(Bertram et al. 2018); Pathway 5: IMAGE (van Vuuren et al. 2018); Pathway 6: MESSAGE-GLOBIOM**
 25 **(Grubler et al. 2018). Data is from an update of the IAMC Scenario Explorer developed for the SR15**
 26 **(Rogelj et al. 2018). The categories CO₂ Land, CH₄ Land and N₂O Land include GHG emissions from**
 27 **land-use change and agricultural land use (including emissions related to bioenergy production). In**
 28 **addition, the category CO₂ Land includes negative emissions due to afforestation. BECCS reflects the**
 29 **CO₂ emissions captured from bioenergy use and stored in geological deposits. Solid lines show the net**
 30 **effect of all land-based GHG emissions and removals (CO₂ Land, CH₄ Land, N₂O Land and BECCS),**
 31 **while dashed lines show the net effect excluding BECCS. CH₄ and N₂O emissions are converted to CO₂-**
 32 **eq using GWP factors of 28 and 265 respectively.**
 33

2.6.3 The contribution of response options to the Paris Agreement

The previous sections indicated how land based response options have the potential to contribute to the Paris Agreement, not only through reducing anthropogenic emissions but also for providing anthropogenic sinks that can contribute to “...a balance between anthropogenic emissions by sources and removals by sinks of greenhouse gases in the second half of this century ...” (Paris Agreement, Article 4). The balance applies globally, and relates only to greenhouse gases, not aerosols (Section 2.4) or biophysical effects (Section 2.5).

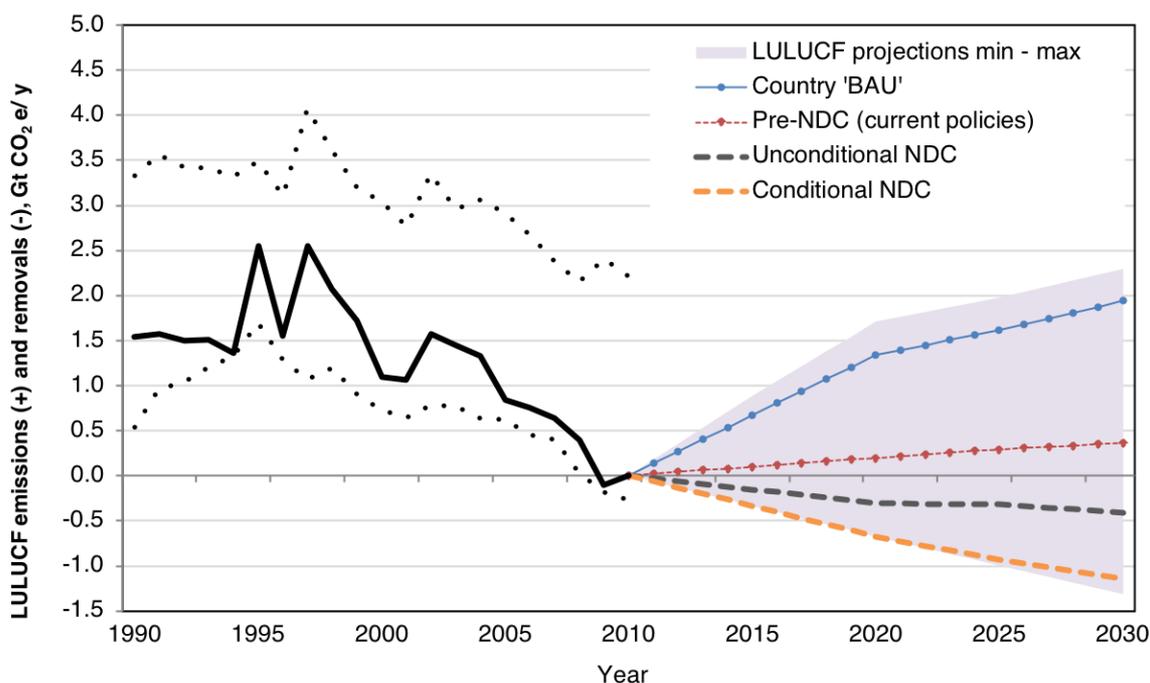
The Paris Agreement includes an Enhanced Transparency Framework, to track countries’ progress towards achieving their individual targets (i.e., NDCs), and a Global Stocktake (every five years starting in 2023), to assess the countries’ collective progress towards the long-term goals of the Paris Agreement. The importance of robust and transparent definitions and methods (including the approach to separating anthropogenic from natural fluxes) (Fuglestad et al. 2018) and the needs for reconciling country GHG inventories and models (Grassi et al. 2018a) was highlighted in 2.3 in relation to estimating emissions. Issues around estimating mitigation is also key to transparency and credibility and is part of the Paris Rulebook.

The land sector is expected to deliver up to 25% of GHG mitigation pledged by countries by 2025-2030 in their NDCs, based on early assessments of “Intended” NDCs submitted ahead of the Paris Agreement and updates immediately after (*low confidence*) (Grassi et al. 2017; Forsell et al. 2016). While most NDCs submitted to date include commitments related to the land sector, they vary with how much information is given and the type of target, with more ambitious targets for developing countries often being “conditional” on support and climate finance. Some do not specify the role of AFOLU but include it implicitly as part of economy-wide pledges (e.g. reducing total emission or emission intensity), a few mention multi-sectoral mitigation targets which include AFOLU in a fairly unspecified manner. Many NDCs include specific AFOLU response options, with most focused on the role of forests, a few included soil carbon sequestration or agricultural mitigation, few explicitly mention bioenergy (e.g., Cambodia, Indonesia and Malaysia), but this could be implicitly included with reduced emission in energy sector through fuel substitution (see Cross-Chapter Box 7: Bioenergy and BECCS in mitigation scenarios, Chapter 6, for discussion on cross sector flux reporting). The countries indicating AFOLU mitigation most prominently were Brazil and Indonesia, followed by other countries focusing either on avoiding carbon emissions (e.g., Ethiopia, Gabon, Mexico, DRC, Guyana and Madagascar) or on promoting the sink through large afforestation programs (e.g., China, India) (Grassi et al. 2017).

Figure 2.28 shows the CO₂ mitigation potential of NDCs compared to historical fluxes from LULUCF⁴. It shows future fluxes based on current policies in place and on country-stated Business As Usual (BAU) activities (these are different from current policies as many countries are already implementing policies that they do not include as part of their historical business-as-usual baseline) (Grassi et al. 2017). Under implementation of unconditional pledges, the net LULUCF flux in 2030 has been estimated to be a sink of $-0.41 \pm 0.68 \text{ GtCO}_2 \text{ yr}^{-1}$, which increases to $-1.14 \pm 0.48 \text{ GtCO}_2 \text{ yr}^{-1}$ in 2030 with conditional activities. This compares to net LULUCF in 2010 calculated from the GHG Inventories of $0.01 \pm 0.86 \text{ GtCO}_2 \text{ yr}^{-1}$ (Grassi et al. 2017). Forsell et al. (2016) similarly find a reduction in 2030 compared to 2010 of $0.5 \text{ GtCO}_2 \text{ yr}^{-1}$ (range: 0.2–0.8) by 2020 and $0.9 \text{ GtCO}_2 \text{ yr}^{-1}$ (range: 0.5–1.3) by 2030 for unconditional and conditional cases.

The approach of countries to calculating the LULUCF contribution towards the NDC varies, with implications for comparability and transparency. For example, by following the different approaches used to include LULUCF in country NDCs, (Grassi et al. 2017) found a 3-fold difference in estimated mitigation: $1.2\text{--}1.9 \text{ GtCO}_2\text{-eq yr}^{-1}$ when 2030 expected emissions are compared to 2005 emissions; $0.7\text{--}1.4 \text{ GtCO}_2\text{-eq yr}^{-1}$ when 2030 emissions are compared to reference scenarios based on current policies or $2.3\text{--}3.0 \text{ GtCO}_2\text{-eq yr}^{-1}$ when compared to BAU, and $3.0\text{--}3.8 \text{ GtCO}_2\text{-eq yr}^{-1}$ when based on using each countries’ approach to calculation stated in the NDC (i.e., when based on a mix of country approaches, using either past years or BAU projections as reference)

⁴ FOOTNOTE: CO₂ fluxes due to Land Use, Land-Use Change and Forestry, i.e. not including the part of AFOLU fluxes that are from agriculture



1
2
3 **Figure 2.28 Global Land Use, Land Use Change and Forestry (LULUCF) net greenhouse gas flux for the**
4 **historical period and future scenarios based on analyses of countries' Nationally Determined Contributions**
5 **NDCs. The LULUCF historical data (black solid line) reflect the following countries' documents (in order of**
6 **priority): data submitted to UNFCCC (NDCs⁵, 2015 GHG Inventories⁶, recent National Communications^{7, 8});**
7 **other official countries' documents; FAO-based datasets, i.e. FAO-FRA for forest (Tian et al. 2015) as**
8 **elaborated by (Federici et al. 2015) and FAOSTAT for non-forest land use emissions (FAO 2015a) . The**
9 **future four scenarios reflect official countries' information, mostly INDCs or updated NDCs available at the**
10 **time of the analysis (Feb 2016), complemented by Biennial Update Reports⁹ and National Communications,**
11 **and show: the Business as Usual (BAU) scenario as defined by the country; the trend based on pre-NDC**
12 **levels of activity (current policies in place in countries); and the unconditional NDC and conditional NDC**
13 **scenarios. The shaded area indicates the full range of countries' available projections (min-max), expressing**
14 **the available countries' information on uncertainties beyond the specific scenarios shown. The range of**
15 **historical country datasets (dotted lines) reflects differences between alternative selections of country sources,**
16 **i.e. GHG inventories for developed countries complemented by FAO-based datasets (upper range) or by data**
17 **in National Communications (lower range) for developing countries**
18
19

20 In Exploring the effectiveness of the NDCs, SR15 concluded “*Estimates of global average temperature*
21 *increase are 2.9–3.4°C above preindustrial levels with a greater than 66% probability by 2100* (Roberts et
22 *al. 2006; Rogelj et al. 2016), under a full implementation of unconditional NDCs and a continuation of*
23 *climate action similar to that of the NDCs. In order to achieve 1.5°C or 2°C pathways, this shortfall would*
24 *imply the need for submission (and achievement) of more ambitious NDCs, and plan for a more rapid*
25 *transformation of their national energy, industry, transport, and land use sectors (Peters and Geden 2017;*
26 *Millar et al. 2017; Rogelj et al. 2016).*

⁵ FOOTNOTE: UNFCCC. INDCs as communicated by Parties,
<http://www4.unfccc.int/submissions/indc/Submission%20Pages/submissions.aspx>. (UNFCCC, 2015).

⁶ FOOTNOTE: UNFCCC. Greenhouse Gas Inventories,
http://unfccc.int/national_reports/annex_i_ghg_inventories/national_inventories_submissions/items/8812.php.
(UNFCCC, 2015).

⁷ FOOTNOTE : UNFCCC. National Communications Non-Annex 1, <http://unfccc.int/nationalreports/non-annexinatcom/submittednatcom/items/653.php> (UNFCCC, 2015).

⁸ FOOTNOTE : UNFCCC. National Communications Annex 1,
<http://unfccc.int/nationalreports/annexinatcom/submittednatcom/items/7742.php>; (UNFCCC, 2015).

⁹ FOOTNOTE : UNFCCC. Biennial Update Reports, http://unfccc.int/national_reports/non-annex_i_natcom/reporting_on_climate_change/items/8722.php (UNFCCC, 2015).

1
2 Response options relying on the use of land could provide around a third of the additional mitigation needed
3 in the near term (2030) to close the gap between current policy trajectories based on NDCs and what is
4 required to achieve a 2°C (>66% chance) or 1.5°C (50 to 66% chance) pathway according to the UNEP
5 Emissions Gap Report (Roberts et al. 2006). The report estimates annual reduction potentials in 2030 from
6 agriculture 3.0 (2.3–3.7) GtCO₂-eq yr⁻¹, a combination of “uncertain measures” (biochar, peat-related
7 emission reductions, and demand-side management) 3.7 (2.6–4.8) GtCO₂-eq yr⁻¹; forests 5.3 (4.1–6.5)
8 GtCO₂-eq yr⁻¹, bioenergy 0.9 GtCO₂-eq yr⁻¹, and BECCS 0.3 (0.2 to 0.4) GtCO₂-eq yr⁻¹ (UNEP 2017) Table
9 4.1). These response options account for 35% of potential reduction (or 32% without bioenergy and BECCS)
10 out of a total (all sector) potential of 38 (35–41) GtCO₂-eq yr⁻¹. The potentials estimated in the UNEP
11 Emissions Gap Report are based on the technical potential of individual response options from literature
12 including that presented in Section 2.1. CDR related to land use, while not a substitute for strong action in
13 the energy sector, has the technical potential to balance unavoidable emissions that are difficult to eliminate
14 with current technologies (*high confidence*), with early action avoiding deeper and more rapid action later
15 (*very high confidence*) (Strefler et al. 2018; Elmar et al. 2018, SR15).
16
17
18

2.7 Plant and soil processes underlying land-climate interactions

Projecting future complex interactions between land and climate require Earth system models (ESMs). A growing number of studies suggested that many processes important for interactions between land and climate were missing in the CMIP5-class ESMs and that Dynamic vegetation models (DGVM) used tended to elevate CO₂ emission and removals (*high confidence*) (Busch and Sage 2017; Rogers et al. 2017; Anderegg et al. 2016; Tjoelker 2018; Sulman et al. 2014a; Wieder et al. 2018; Davidson et al. 2006a).

Ecosystem complexity stemming from the diversity of plants, animals and microbes, as well as their biological responses to gradual climate changes (e.g., adaptive migration) and disturbance events (e.g., extreme weather events, fire, pest outbreaks; Section 2.2), are of potential importance. Of these processes, this section focuses on plant and soil processes, as recent empirical work, including those explained in the following subsections, offer potential for improved model projections under warmer and CO₂ rich futures.

The magnitude of future uptake and release of CO₂ and other greenhouse gases by vegetation are among the greatest uncertainties (Ciais et al. 2013b). One reason for this uncertainty stems from the lack of understanding of the mechanisms responsible for plant responses to increasing temperatures. The short- and long-term projections of gross photosynthesis responses to changes in temperature, CO₂, nutrient availability vary greatly among the models (Busch and Sage 2017; Rogers et al. 2017). Net CO₂ exchange requires estimation of autotrophic respiration, which is another source of uncertainty in ESM projections (Malhi et al. 2011). The importance of plant acclimation of photosynthesis and respiration in understanding vegetation response to climate change is now widely recognised (*high confidence*) (Rogers et al., 2017; Tan et al., 2017; Tjoelker, 2018; Vanderwel et al., 2015 ; Section 2.7.1). Acclimation is broadly defined as the biochemical, physiological, morphological or developmental adjustments within the lifetime of organisms that result in improved performance at the new condition. Acclimation often operates over a time span of days to weeks, and can mitigate negative effects of climate change on organismal growth and ecosystem functions (Tjoelker 2018).

Soil carbon and microbial processes, which interact with plant responses to climate, represent another large source of uncertainty in model projections (*medium confidence*) (Sections 2.7.2, 2.7.3 and 2.7.4). Given the wide range of uncertainty associated with SOC size estimates, CMIP5 models use a wide range of starting SOC stocks from 510 to 3040 GtC (Todd-Brown et al. 2013). Soil microbial respiration is estimated to release 40–70 GtC annually from the soil to the atmosphere globally (Hawkes et al. 2017). Projections of changes in global SOC stocks during the 21st century by CMIP5 models also ranged widely, from a loss of 37 Gt to a gain of 146 Gt, with differences largely explained by initial SOC stocks, differing C input rates, and different decomposition rates and temperature sensitivities (Todd-Brown et al. 2013). With respect to land-climate interactions, the key processes affecting SOC stocks are warming (which is expected to accelerate SOC losses through microbial respiration) and acceleration of plant growth (which increases inputs of C to soils). However, complex mechanisms underlying SOC responses to moisture regimes, carbon addition, and warming drive considerable uncertainty in projections of future changes in SOC stocks (Sulman et al. 2014a; Singh et al. 2010; Wieder et al. 2018).

2.7.1 Temperature responses of plant and ecosystem production

Climate-change responses of net ecosystem production cannot be modelled by simple instantaneous response functions, because of thermal acclimation responses of plants and soil microbes, as well as delayed responses arising from interactions between plants and the soil (*high confidence*) (Slot et al. 2014; Rogers et al. 2017; Tan et al. 2017; Tjoelker 2018). Photosynthesis and respiration of component plant species exhibit different functional shapes among species (Slot et al. 2014), and carbon balance at the stand level is influenced by respiration of ecosystem biomass other than plants. Large uncertainty remains for thermal responses of bacteria and other soil organisms (Section 2.7.5). Bayesian statistical estimates of global photosynthesis and total ecosystem respirations suggest that they exhibit different responses to thermal anomaly during the last 35 years (Li et al. 2018b).

Thermal responses of plant respiration, which consumes approximately one half of GPP, have not been

1 appropriately incorporated in most ESMs (Davidson et al., 2006; Tjoelker, 2018). Assumptions associated
2 with respiration have been a major source of uncertainty for ESMs at the time of AR5. In most existing
3 models, a simple assumption that respiration doubles with each 10°C increase of temperature (i.e., $Q_{10} = 2$) is
4 adopted, ignoring acclimation. Even a small error stemming from this assumption can strongly influence
5 estimated net carbon balance at large spatial scales of ecosystems and biomes over the time period of
6 multiple decades (Smith and Dukes 2013; Smith et al. 2016b). In order to estimate more appropriate thermal
7 response curves of respiration, a global database including data from 899 plant species has been compiled
8 (Atkin et al. 2015), and respiration data from 231 plants species across seven biomes have been analysed
9 (Heskel et al. 2016). These empirical data on thermal responses of respiration demonstrate a globally
10 convergent pattern (Huntingford et al. 2017). According to a sensitivity analysis of a relatively small number
11 of ESMs, a newly derived function of instantaneous responses of plant respiration to temperature (instead of
12 a traditional exponential function of $Q_{10} = 2$) makes a significant difference in estimated autotrophic
13 respiration especially in cold biomes (Heskel et al. 2016).

14
15 Acclimation results in reduced sensitivity of plant respiration with rising temperature, i.e., down regulation
16 of warming-related increase in respiratory carbon emission (Atkin et al. 2015; Slot and Kitajima 2015;
17 Tjoelker 2018) (*high confidence*). For example, experimental data from a tropical forest canopy show that
18 temperature acclimation ameliorates the negative effects of rising temperature to leaf and plant carbon
19 balance (Slot et al. 2014). Analysis of CO₂ flux data to quantify optimal temperature of net primary
20 production of tropical forests also suggest acclimation potential for many tropical forests (Tan et al. 2017).
21 Comparisons of models with and without thermal acclimation of respiration show that acclimation can halve
22 the increase of plant respiration with projected temperature increase by the end of 21st century (Vanderwel et
23 al. 2015).

24
25 It is typical that acclimation response to warming results in increases of the optimum temperature for
26 photosynthesis and growth (Slot and Winter 2017; Yamori et al. 2014; Rogers et al. 2017). Although such
27 shift is a result of a complex interactions of biochemical, respiratory, and stomatal regulation (Lloyd and
28 Farquhar 2008), it can be approximated by a simple algorithm to address acclimation (Kattge et al. 2007).
29 Mercado et al., (2018), using this approach, found that inclusion of biogeographical variation in
30 photosynthetic temperature response was critically important for estimating future land surface carbon
31 uptake. In the tropics, CO₂ fertilisation effect (c.f., Box 2.3) is suggested to be more important for observed
32 increases in carbon sink strength than increased leaf area index or longer growing season (Zhu et al. 2016).
33 Acclimation responses of photosynthesis and growth to simultaneous changes of temperature and CO₂, as
34 well as stress responses above the optimal temperature for photosynthesis, remain a major knowledge gap in
35 modelling responses of plant productivity under future climate change (Rogers et al. 2017).

36 37 **2.7.2 Water transport through soil-plant-atmosphere continuum and drought mortality**

38
39 How climate change, especially changes of precipitation patterns, influence water transport through the soil-
40 plant-atmosphere continuum, is a key element in projecting the future of water vapour flux from land and
41 cooling via latent heat flux (Sellers et al. 1996; Bonan 2008a; Brodribb 2009; Choat et al. 2012; Sperry and
42 Love 2015; Novick et al. 2016; Sulman et al. 2016)(*high confidence*). Even without changes in leaf area per
43 unit area of land, when plants close stomata in response to water shortage, dry atmosphere, or soil moisture
44 deficit, the stand-level fluxes of water (and associated latent heat flux) decrease (Seneviratne et al. 2018).
45 Closing stomata enhances drought survival at the cost of reduced photosynthetic production, while not
46 closing stomata avoids loss of photosynthetic production at the cost of increased drought mortality (Sperry
47 and Love 2015). Hence, species-specific responses to drought, in terms of whether they close stomata or not,
48 have short and long-term consequences (Anderegg et al. 2018a; Buotte et al. 2019). Increased drought-
49 induced mortality of forest trees, often exacerbated by insect outbreak and fire (e.g., (Breshears et al. 2005;
50 Kurz et al. 2008a; Allen et al. 2010)) (Section 2.2.4), have long-term impact on hydrological interactions
51 between land and atmosphere (Anderegg et al. 2018b).

52
53 New models linking plant water transport with canopy gas exchange and energy fluxes are expected to
54 improved projections of climate change impacts on forests and land-atmosphere interactions (Bohrer et al.,
55 2005; Anderegg et al., 2016; Sperry and Love, 2015; Wolf et al., 2016)(*medium confidence*). Yet, there is
56 much uncertainty in the ability of current vegetation and land surface models to adequately capture tree

1 mortality and the response of forests to climate extremes like drought (Rogers et al. 2017; Hartmann et al.
2 2018). Most vegetation models use climate stress envelopes or vegetation carbon balance estimations to
3 project climate-driven mortality and loss of forests (McDowell et al. 2011); these may not adequately project
4 biome shifts and impacts of disturbance in future climates. For example, a suite of vegetation models was
5 compared to a field drought experiment in the Amazon on mature rainforest trees and all models performed
6 poorly in projecting the timing and magnitude of biomass loss due to drought (Powell et al. 2013). More
7 recently, the loss of water transport due to embolism (disruption of xylem water continuity) (Sperry and
8 Love 2015), rather than carbon starvation (Rowland et al. 2015), is receiving attention as a key physiological
9 process relevant for drought-induced tree mortality (Hartmann et al. 2018). A key challenge to modelling
10 effort is to consider differences among plant species and vegetation types in their drought responses. One
11 approach is to classify plant species to “functional types” that exhibit similar responses to environmental
12 variations (Anderegg et al. 2016). Certain traits of species, such as tree height, is shown to be predictive of
13 growth decline and mortality in response to drought (Xu et al. 2016a). Similarly, tree rooting depth is
14 positively related to mortality, contrary to expectation, during prolonged droughts in tropical dry forest
15 (Chitra-Tarak et al. 2017).

16 2.7.3 Soil microbial effects on soil nutrient dynamics and plant responses to elevated CO₂

17
18
19 Soil microbial processes influencing nutrient and carbon dynamics represent a large source of uncertainty in
20 projecting land-climate interactions. For example, ESMs incorporating nitrogen and phosphorus limitations
21 (but without considering the effects of mycorrhizae and rhizosphere priming) indicate that the simulated
22 future C-uptake on land is reduced significantly when both nitrogen and phosphorus are limited as compared
23 to only C-stimulation, by 63% (of 197 Pg C) under RCP2.6 and by 67% (of 425 Pg C) under RCP8.5 (Zhang
24 et al. 2013c). Mineral nutrient limitation progressively reduces the CO₂ fertilisation effects on plant growth
25 and productivity over time (Norby et al. 2010; Sardans et al. 2012; Reich and Hobbie 2013; Feng et al. 2015;
26 Terrer et al. 2017) (*robust evidence, medium agreement*). The rates at which nutrient limitation develops
27 differ among studies and sites. A recent meta-analysis shows that experimental CO₂ enrichment generally
28 results in lower nitrogen and phosphorus concentrations in plant tissues (Du et al. 2019a), and isotopic
29 analysis also suggest a global trend of decreases in leaf nutrient concentration (Craine et al. 2018; Jonard et
30 al. 2015). However, reduced responses to elevated CO₂ (eCO₂) may not be a simple function of nitrogen
31 dilution per se, as they result from complex interactions of ecosystem factors that influence nitrogen
32 acquisition by plants (Liang et al. 2016; Rutting 2017; Du et al. 2019a).

33
34 Increasing number of case studies suggest that soil microbial processes, such as nitrogen mineralisation rates
35 and symbiosis with plants, influence nutrient limitation on eCO₂ effects on plant growth (Drake et al. 2011;
36 Zak et al. 2011; Hungate et al. 2013; Talhelm et al. 2014; Du et al. 2019a) (*medium confidence*).
37 Rhizosphere priming effects (i.e., release of organic matters by roots to stimulate microbial activities) and
38 mycorrhizal associations are proposed to explain why some sites becoming nitrogen limited after a few years
39 and others sustaining growth through accelerated nitrogen uptake (Phillips et al. 2011; Terrer et al. 2017)
40 (*limited evidence, medium agreement*).

41
42 Model assessments that including rhizosphere priming effects and ectomycorrhizal symbiosis suggest that
43 soil organic matter (SOM) cycling is accelerated through microbial symbiosis (Elbert et al. 2012; Sulman et
44 al. 2017; Orwin et al. 2011; Baskaran et al. 2017) (*medium confidence*). Uncertainty exists in differences
45 among ectomycorrhizal fungal species in their ability to decompose SOM (Pellitier and Zak 2018) and the
46 capacity of ecosystems to sustain long-term growth with these positive symbiotic feedbacks is still under
47 debate (Terrer et al. 2017). ESMs include only biological nitrogen cycles, even though a recent study
48 suggests that bedrock weathering can be a significant source of nitrogen to plants (Houlton et al. 2018). In
49 contrast, rock weathering is widely considered to be the key for P availability, and tropical forests with
50 highly weathered soils are considered to be limited by P availability rather than nitrogen availability (Reed et
51 al. 2015). Yet, evidence from P-fertilisation experiments is lacking (Schulte-Uebbing and de Vries 2018) and
52 P limitation of tropical tree growth may be strongly species-specific (Ellsworth et al. 2017; Turner et al.
53 2018a). Limitation by availability of soil nutrients other than nitrogen and P has not been studied in the
54 context of land-climate interactions, except potassium (K) as a potentially limiting factor for terrestrial plant
55 productivity in interaction with N, P and hydrology (Sardans and Peñuelas 2015; Zhao et al. 2017; Wright et
56 al. 2018).

1
2 Anthropogenic alteration of global and regional nitrogen and P cycles, largely through use of chemical
3 fertilisers and pollution, has major implications for future ecosystem attributes, including C storage, in
4 natural and managed ecosystems (Peñuelas et al. 2013, 2017; Wang et al. 2017c; Schulte-Uebbing and de
5 Vries 2018; Yuan et al. 2018) (*high confidence*). During 1997-2013, the contribution of nitrogen deposition
6 to the global C sink has been estimated at $0.27 (\pm 0.13)$ GtC yr⁻¹, and the contribution of P deposition as
7 $0.054 (\pm 0.10)$ GtC yr⁻¹; these constitute about 9% and 2% of the total land C sink, respectively (Wang et al.
8 2017c). Anthropogenic deposition of nitrogen enhances carbon sequestration by vegetation (Schulte-
9 Uebbing and de Vries 2018), but this effect of nitrogen deposition on carbon sequestration may be offset by
10 increased emission of GHGs such as N₂O and CH₄ (Liu and Greaver 2009). Furthermore, nitrogen deposition
11 may lead to imbalance of nitrogen vs. phosphorus availability (Peñuelas et al. 2013), soil microbial activity
12 and SOM decomposition (Janssens et al. 2010), and reduced ecosystem stability (Chen et al. 2016b).

13 14 **2.7.4 Vertical distribution of soil organic carbon**

15
16 It has long been recognised that dynamics of soil organic carbon (SOC) represent a large source of
17 uncertainties on biogeochemical interactions of land with atmosphere and climate as detailed below. Since
18 AR5, there have been new understanding on SOC size, as well as microbial processes that influence SOM
19 dynamics under climate change and LULCC. Three existing data bases (SoilGrids, the Harmonized World
20 Soil Data Base, Northern Circumpolar Soil Database) substantially differ in estimated size of global soil
21 organic carbon (SOC) stock down to 1 m depth, varying between 2500 Pg to 3400 Pg with differences
22 among databases largely attributable to C stored in permafrost (Joosten 2015; Köchy et al. 2015; Tifafi et al.
23 2018). These values are four to eight times larger than the carbon stock associated with the terrestrial
24 vegetation (Bond-Lamberty et al. 2018). New estimates since AR5 show that much larger areas in Amazon
25 and Congo basins are peatlands (Gumbrecht et al. 2017; Dargie et al. 2019).

26
27 Deep soil layers can contain much more carbon than previously assumed (e.g., González-Jaramillo et al.,
28 2016) (*limited evidence, medium agreement*). Based on radiocarbon measurements, deep SOC can be very
29 old, with residence times up to several thousand years (Rumpel and Kögel-Knabner 2011) or even several
30 tens of thousands of years (Okuno and Nakamura 2003). Dynamics associated with such deeply buried
31 carbon remain poorly studied and ignored by the models, and not addressed in most of the studies assessed in
32 this subsection. Deep soil C is thought to be stabilised by mineral interactions, but recent experiments
33 suggest that CO₂ release from deep soils can also be increased by warming, with a 4°C warming enhancing
34 annual soil respiration by 34–37% (Hicks Pries et al. 2017) or addition of fresh carbon (Sebastien
35 Fontaine Pierre Barre, Nadia Bdioui, Bruno Mary, Cornelia Rumpel 2007). While erosion is not typically
36 modelled as a carbon flux in ESMs, erosion and burial of carbon-containing sediments is likely a significant
37 carbon transfer from land to ocean (Berhe et al. 2007; Asefaw et al. 2008; Wang et al. 2017e) (*medium
38 confidence*).

39 40 **2.7.5 Soil carbon responses to warming and changes in soil moisture**

41
42 Annually, 119 GtC is estimated to be emitted from the terrestrial ecosystem to the atmosphere, of which ca.
43 50% is attributed to soil microbial respiration (Auffret et al. 2016; Shao et al. 2013). It is yet not possible yet
44 to make mechanistic and quantitative projections about how multiple environmental factors influence soil
45 microbial respiration (Davidson et al. 2006a; Dungait et al. 2012). Soil warming experiments show
46 significant variability in temperature and moisture responses across biomes and climates; Crowther et al.,
47 (2016) found that warming-induced SOC loss is greater in regions with high initial carbon stocks, while an
48 analysis of an expanded version of the same dataset did not support this conclusion (Gestel et al. 2018).
49 Studies of SOC responses to warming over time have also shown complex responses. In a multi-decadal
50 warming experiment, Melillo et al., (2017) found that soil respiration response to warming went through
51 multiple phases of increasing and decreasing strength, which were related to changes in microbial
52 communities and available substrates over time. (Conant et al., (2011) and Knorr et al., (2005) suggested that
53 transient decomposition responses to warming could be explained by depletion of labile substrates, but that
54 long-term SOC losses could be amplified by high temperature sensitivity of slowly decomposing SOC
55 components. Overall, long-term SOC responses to warming remain uncertain (Davidson et al. 2006a;
56 Dungait et al. 2012; Nishina et al. 2014; Tian et al. 2015).

1
2 It is widely known that soil moisture plays an important role in SOM decomposition by influencing
3 microbial processes (e.g., Monard et al., 2012; Moyano et al., 2013; Yan et al., 2018), as confirmed by a
4 recent global meta-analysis (Hawkes et al. 2017) (*high confidence*). A likely mechanism involves that
5 increased soil moisture lowers C mineralisation rates under anaerobic conditions resulting in enhanced C
6 stocks, but experimental analyses have shown that this effect may last for only 3–4 weeks after which iron
7 reduction can actually accelerate loss of previously protected organic C by facilitating microbial access
8 (Huang and Hall 2017).

9
10 Experimental studies of responses of microbial respiration to warming have found variable results (Luo et al.
11 2001; Bradford et al. 2008; Zhou et al. 2011; Carey et al. 2016; Teramoto et al. 2016). No acclimation was
12 observed in C-rich calcareous temperate forest soils (Schindlbacher et al. 2015) and arctic soils (Hartley et
13 al. 2008), and a variety of ecosystems from the Arctic to the Amazon indicated that microbes appear to
14 enhance the temperature sensitivity of soil respiration in Arctic and boreal soils, thereby releasing even more
15 carbon than currently projected (Karhu et al. 2014). In tropical forests, P limitation of microbial processes is
16 a key factor influencing soil respiration (Camenzind et al. 2018). Temperature responses of symbiotic
17 mycorrhizae differ widely among host plant species, without a clear pattern that may allow generalisation
18 across plant species and vegetation types (Fahey et al. 2016).

19
20 Some new insights have been obtained since AR5 from investigations of improved mechanistic
21 understanding of factors that regulate temperature responses of soil microbial respiration. Carbon use
22 efficiency and soil nitrogen dynamics have large influence on SOC responses to warming (Allison et al.
23 2010; Frey et al. 2013; Wieder, William R., Bonan, Gordon B., Allison 2013; García-Palacios et al. 2015)
24 (*high confidence*). More complex community interactions including competitive and trophic interactions
25 could drive unexpected responses to SOC cycling to changes in temperature, moisture, and C inputs
26 (Crowther et al. 2015; Buchkowski et al. 2017). Competition for nitrogen among bacteria and fungi could
27 also suppress decomposition (Averill et al. 2014). Overall, the roles of soil microbial community and trophic
28 dynamics in global SOC cycling remain very uncertain.

29 30 **2.7.6 Soil carbon responses to changes in organic-matter inputs by plants**

31
32 While current ESM structures mean that increasing C inputs to soils drive corresponding increases in SOC
33 stocks, long-term carbon addition experiments have found contradictory SOC responses. Some litter addition
34 experiments have observed increased SOC accumulation (Lajtha et al. 2014b; Liu et al. 2009), while others
35 suggest insignificant SOC responses (Lajtha et al. 2014a; van Groenigen et al. 2014). Microbial dynamics
36 are believed to have an important role in driving complex responses to C additions. The addition of fresh
37 organic material can accelerate microbial growth and SOM decomposition via priming effects (Kuzyakov et
38 al. 2014; Cheng et al. 2017). SOM cycling is dominated by “hot spots” including the rhizosphere as well as
39 areas surrounding fresh detritus (*medium evidence, high agreement*) (Finzi et al. 2015; Kuzyakov and
40 Blagodatskaya 2015). This complicates projections of SOC responses to increasing plant productivity;
41 increasing C inputs could promote higher SOC storage, but these fresh C inputs could also deplete SOC
42 stocks by promoting faster decomposition (Hopkins et al. 2014; Guenet et al. 2018; Sulman et al. 2014b). A
43 meta-analysis by (van Groenigen et al. 2014) suggested that elevated CO₂ accelerated SOC turnover rates
44 across several biomes. These effects could be especially important in high-latitude regions where soils have
45 high organic matter content and plant productivity is increasing (Hartley et al. 2012), but have also been
46 observed in the tropics (Sayer et al. 2011).

47
48 Along with biological decomposition, another source of uncertainty in projecting responses of SOC to
49 climate change is stabilisation via interactions with mineral particles (Kögel-Knabner et al. 2008; Kleber et
50 al. 2011; Marschner et al. 2008; Schmidt 2011) (*high confidence*). Historically, conceptual models of SOC
51 cycling have centred on the role of chemical recalcitrance, the hypothesis that long-lived components of
52 SOC are formed from organic compounds that are inherently resistant to decomposition. Under the emerging
53 new paradigm, stable SOC is primarily formed by the bonding of microbially-processed organic material to
54 mineral particles, which limits the accessibility of organic material to microbial decomposers (Lützow et al.
55 2006; Keiluweit et al. 2015; Kallenbach et al. 2016; Kleber et al. 2011; Hopkins et al. 2014). SOC in soil
56 aggregates can be protected from microbial decomposition by being trapped in soil pores too small for

1 microbes to access (Blanco-Canqui and Lal 2004; Six et al. 2004) or by oxygen limitation (Keiluweit et al.
2 2016). Some new models are integrating these mineral protection processes into SOC cycling projections
3 (Wang et al. 2017a; Sulman et al. 2014b; Riley et al. 2014; Wieder et al. 2015), although the sensitivity of
4 mineral-associated organic matter to changes in temperature, moisture, fire (see Box 2.1) and carbon inputs
5 is highly uncertain. Improved quantitative understanding of soil ecosystem processes will be critically
6 important for projection of future land-climate feedback interactions.
7
8

9 **Frequently Asked Questions**

10 **FAQ 2.1 How does climate change affect land use and land cover?**

11
12
13 Contemporary land cover and land use is adapted to current climate variability within particular temperature
14 and/or rainfall ranges (referred to as climate envelopes). Anthropogenic greenhouse gas emissions impact
15 land through changes in the weather and climate and also through modifications in atmospheric composition
16 through increased greenhouse gasses, especially CO₂. A warming climate alters the current regional climate
17 variability and results in a shift of regional climate envelopes poleward and to higher elevations. The shift of
18 warmer climate envelopes into high latitude areas has potential benefits for agriculture here through
19 extended growing seasons and warmer seasonal temperatures and increased atmospheric CO₂ concentrations
20 enhances photosynthetic activity. However, this warming will also lead to enhanced snowmelt and reduced
21 albedo, permafrost melting and the further release of methane and CO₂ into the atmosphere as the permafrost
22 begins to decompose. Concurrent with these climate envelope shifts will be the emergence of new, hot
23 climates in the tropics and increases in the frequency, intensity and duration of extreme events (e.g. heat
24 waves, very heavy rainfall, drought). These emergent hot climates will negatively affect land use (through
25 changes in crop productivity, irrigation needs, management practices) and land cover through loss of
26 vegetation productivity in many parts of the world, and overwhelm any benefits to land use and land cover
27 derived from increased atmospheric CO₂ concentrations.
28

29 **FAQ 2.2 How do the land and land use contribute to climate change?**

30
31 Any changes to the land and how it is used can affect exchanges of water, energy, greenhouse gases (e.g.,
32 CO₂, CH₄, N₂O), non-greenhouse gases (e.g., biogenic volatile organic compounds – BVOCs), and aerosols
33 (mineral, e.g., dust, or, carbonaceous, e.g., black carbon) between the land and the atmosphere. Land and
34 land use change therefore alter the state (e.g., chemical composition and air quality, temperature and
35 humidity) and the dynamics (e.g., strength of horizontal and vertical winds) of the atmosphere, which in turn
36 can dampen or amplify local climate change. Land-induced changes in energy, moisture and wind can affect
37 neighbouring, and sometimes more distant, areas. For example, deforestation in Brazil warms the surface, in
38 addition to global warming, and enhances convection which increases the relative temperature difference
39 between the land and the ocean, boosting moisture advection from the ocean and thus rainfall further inland.
40 Vegetation absorbs carbon dioxide (CO₂) to use for growth and maintenance. Forests contain more carbon in
41 their biomass and soils than croplands and so a conversion of forest to cropland, for example, results in
42 emissions of CO₂ to the atmosphere, thereby enhancing the greenhouse gas-induced global warming.
43 Terrestrial ecosystems are both sources and sinks of chemical compounds such as nitrogen and ozone.
44 BVOCs contribute to forming tropospheric ozone and secondary aerosols, which respectively affect surface
45 warming and cloud formation. Semi-arid and arid regions release dust, as do cropland areas after harvest.
46 Increasing the amount of aerosols in the atmosphere impacts temperature in both positive and negative ways
47 depending on the particle size, altitude, and nature (carbonaceous or mineral for example). Although global
48 warming will impact the functioning and state of the land (see FAQ 2.1), this is not a one-way interaction as
49 changes in land and land use can also affect climate and thus modulate climate change. Understanding this
50 two-way interaction can help improve adaptation and mitigation strategies as well as manage landscapes.
51

52 **FAQ 2.3 How does climate change affect water resources?**

1
2 Renewable freshwater resources are essential for the survival of terrestrial and aquatic ecosystems and
3 human use in agriculture, industry and domestically. As increased water vapour concentrations are expected
4 in a warmer atmosphere, climate change will alter the hydrological cycle and therefore regional freshwater
5 resources. In general, wet regions are projected to get wetter and dry regions drier, although there are
6 regional exceptions to this. The consequent impacts vary regionally; where rainfall is projected to be lower
7 in the future (many arid subtropical regions and those with a Mediterranean climate), a reduction of water
8 resources is expected. Here increased temperatures and decreased rainfall will reduce surface and
9 groundwater resources, increase plant evapotranspiration and increase evaporation rates from open water
10 (rivers, lakes, wetlands) and water supply infrastructure (canals, reservoirs). In regions where rainfall is
11 projected to be higher in the future (many high latitude regions and the wet tropics), an increase in water
12 resources can be expected to benefit terrestrial and freshwater ecosystems, agriculture and domestic use,
13 however, these benefits may be limited due to increased temperatures. An increase in extreme rainfall events
14 is also expected which will lead to increases in surface runoff, regional flooding and nutrient removal as well
15 as a reduction in soil water and groundwater recharge in many places. Anthropogenic land use change may
16 amplify or moderate the climate change effect on water resources therefore informed land management
17 strategies need to be developed. A warming climate will exacerbate the existing pressures on renewable
18 freshwater resources in water-stressed regions of the Earth and result in increased competition for water
19 between human and natural systems.
20
21

References

- 1 Abatzoglou, J. T., and A. P. Williams, 2016: Impact of anthropogenic climate change on wildfire across
2 western US forests. *Proc. Natl. Acad. Sci.*, **113**, 11770–11775, doi:10.1073/pnas.1607171113.
- 3 Abdalla, M., A. Hastings, J. Truu, M. Espenberg, Ü. Mander, and P. Smith, 2016: Emissions of methane
4 from northern peatlands: a review of management impacts and implications for future management
5 options. *Ecol. Evol.*, **6**, 7080–7102, doi:10.1002/ece3.2469.
- 6 Abiodun, B. J., Z. D. Adeyewa, P. G. Oguntunde, A. T. Salami, and V. O. Ajayi, 2012: Modeling the
7 impacts of reforestation on future climate in West Africa. *Theor. Appl. Climatol.*, **110**, 77–96,
8 doi:10.1007/s00704-012-0614-1.
- 9 Achard, F., and Coauthors, 2014: Determination of tropical deforestation rates and related carbon losses
10 from 1990 to 2010. *Glob. Chang. Biol.*, **20**, 2540–2554, doi:10.1111/gcb.12605.
- 11 Adger, W. N., N. W. Arnell, R. Black, S. Dercon, A. Geddes, and D. S. G. Thomas, 2015: Focus on
12 environmental risks and migration: Causes and consequences. *Environ. Res. Lett.*, **10**,
13 doi:10.1088/1748-9326/10/6/060201.
- 14 Adnan, S., K. Ullah, S. Gao, A. H. Khosa, and Z. Wang, 2017: Shifting of agro-climatic zones, their drought
15 vulnerability, and precipitation and temperature trends in Pakistan. *Int. J. Climatol.*, **37**, 529–543,
16 doi:10.1002/joc.5019. <http://doi.wiley.com/10.1002/joc.5019> (Accessed April 15, 2019).
- 17 Ahlgren, S., and L. Di Lucia, 2014: Indirect land use changes of biofuel production - A review of modelling
18 efforts and policy developments in the European Union. *Biotechnol. Biofuels*, doi:10.1186/1754-6834-
19 7-35.
- 20 Ahlswede, B. J., and R. Q. Thomas, 2017: Community earth system model simulations reveal the relative
21 importance of afforestation and forest management to surface temperature in Eastern North America.
22 *Forests*, **8**, 1–10, doi:10.3390/f8120499.
- 23 Ainsworth, E. A., and A. Rogers, 2007: The response of photosynthesis and stomatal conductance to rising
24 [CO₂]: Mechanisms and environmental interactions. *Plant, Cell Environ.*, **30**, 258–270,
25 doi:10.1111/j.1365-3040.2007.01641.x.
- 26 Akagi, S. K., R. J. Yokelson, C. Wiedinmyer, M. J. Alvarado, J. S. Reid, T. Karl, J. D. Crouse, and P. O.
27 Wennberg, 2011: Emission factors for open and domestic biomass burning for use in atmospheric
28 models. *Atmos. Chem. Phys.*, **11**, 4039–4072, doi:10.5194/acp-11-4039-2011.
- 29 Alcántara, V., A. Don, R. Well, and R. Nieder, 2016: Deep ploughing increases agricultural soil organic
30 matter stocks. *Glob. Chang. Biol.*, doi:10.1111/gcb.13289.
- 31 Aldersley, A., S. J. Murray, and S. E. Cornell, 2011: Global and regional analysis of climate and human
32 drivers of wildfire. *Sci. Total Environ.*, **409**, 3472–3481, doi:10.1016/j.scitotenv.2011.05.032.
- 33 Alfieri, L., L. Feyen, and G. Di Baldassarre, 2016: Increasing flood risk under climate change: a pan-
34 European assessment of the benefits of four adaptation strategies. *Clim. Change*, **136**, 507–521,
35 doi:10.1007/s10584-016-1641-1. <https://doi.org/10.1007/s10584-016-1641-1>.
- 36 ———, B. Bisselink, F. Dottori, G. Naumann, A. de Roo, P. Salamon, K. Wyser, and L. Feyen, 2017: Global
37 projections of river flood risk in a warmer world. *Earth's Futur.*, **5**, 171–182,
38 doi:10.1002/2016EF000485.
- 39 Alghamdi, A. S., and T. W. Moore, 2015: Detecting Temporal Changes in Riyadh's Urban Heat Island. *Pap.*
40 *Appl. Geogr.*, **1**, 312–325, doi:10.1080/23754931.2015.1084525.
41 <http://www.tandfonline.com/doi/full/10.1080/23754931.2015.1084525>.
- 42 Alizadeh-Choobari, O., P. Ghafarian, and P. Adibi, 2016: Inter-annual variations and trends of the urban
43 warming in Tehran. *Atmos. Res.*, **170**, 176–185, doi:10.1016/j.atmosres.2015.12.001.
- 44 Alkama, R., and A. Cescatti, 2016: Biophysical climate impacts of recent changes in global forest cover.
45 *Science (80-.)*, **351**, 600–604, doi:10.1126/science.aac8083.
- 46 Allan, J. D., and Coauthors, 2014: Airborne observations of IEPOX-derived isoprene SOA in the Amazon
47 during SAMBBA. *Atmos. Chem. Phys.*, **14**, 11393–11407, doi:10.5194/acp-14-11393-2014.
- 48 Allan, R. P., and C. Liu, 2018: Evaluating Large-Scale Variability and Change in Tropical Rainfall and Its
49 Extremes. *Tropical Extremes*, V. Venugopal, J. Sukhatme, R. Murtugudde, and R. Roca, Eds., Elsevier,
50 139–163.
- 51 Allen, C. D., and Coauthors, 2010: A global overview of drought and heat-induced tree mortality reveals
52 emerging climate change risks for forests. *For. Ecol. Manage.*, **259**, 660–684,
53 doi:10.1016/j.foreco.2009.09.001.
54 <http://www.sciencedirect.com/science/article/pii/S037811270900615X>.
- 55 Allen, M., and Coauthors, 2018: Framing and Context. *Global Warming of 1.5 °C an IPCC special report on*
56

- 1 *the impacts of global warming of 1.5 °C above pre-industrial levels and related global greenhouse gas*
2 *emission pathways, in the context of strengthening the global response to the threat of climate change,*
3 V. Masson-Delmotte et al., Eds. <http://www.ipcc.ch/report/sr15/>.
- 4 Allen, R. J., T. Hassan, C. A. Randles, and H. Su, 2019: Enhanced land–sea warming contrast elevates
5 aerosol pollution in a warmer world. *Nat. Clim. Chang.*, doi:10.1038/s41558-019-0401-4.
- 6 Allison, S. D., M. D. Wallenstein, and M. A. Bradford, 2010: Soil-carbon response to warming dependent on
7 microbial physiology. *Nat. Geosci.*, **3**, 336–340, doi:10.1038/ngeo846.
8 <http://dx.doi.org/10.1038/ngeo846>.
- 9 Alter, R. E., E. S. Im, and E. A. B. Eltahir, 2015: Rainfall consistently enhanced around the Gezira Scheme
10 in East Africa due to irrigation. *Nat. Geosci.*, **8**, 763–767, doi:10.1038/ngeo2514.
- 11 Álvaro-Fuentes, J., M. V López Sánchez, C. Cantero-Martínez, and J. L. Arrúe Ugarte, 2008: Tillage effects
12 on soil organic carbon fractions in Mediterranean dryland agroecosystems. *Soil Sci. Soc. Am. J.*, **72**,
13 541–547, doi:10.2136/sssaj2007.0164.
- 14 Anav, A., P. M. Ruti, V. Artale, and R. Valentini, 2010: Modelling the effects of land-cover changes on
15 surface climate in the Mediterranean region. *Clim. Res.*, **41**, 91–104, doi:10.3354/cr00841.
- 16 —, and Coauthors, 2013: Evaluating the Land and Ocean Components of the Global Carbon Cycle in the
17 CMIP5 Earth System Models. *J. Clim.*, **26**, 6801–6843, doi:10.1175/JCLI-D-12-00417.1.
18 <http://journals.ametsoc.org/doi/abs/10.1175/JCLI-D-12-00417.1>.
- 19 Andela, N., and Coauthors, 2017: A human-driven decline in global burned area. *Science (80-.)*, **356**, 1356–
20 1362, doi:10.1126/science.aal4108. <http://science.sciencemag.org/>.
- 21 Anderegg, W. R. L., T. Klein, M. Bartlett, L. Sack, A. F. A. Pellegrini, B. Choat, and S. Jansen, 2016: Meta-
22 analysis reveals that hydraulic traits explain cross-species patterns of drought-induced tree mortality
23 across the globe. *Proc. Natl. Acad. Sci.*, **113**, 5024–5029, doi:10.1073/pnas.1525678113.
24 <http://www.pnas.org/lookup/doi/10.1073/pnas.1525678113>.
- 25 —, and Coauthors, 2018a: Woody plants optimise stomatal behaviour relative to hydraulic risk. *Ecol.*
26 *Let.*, **0**, doi:10.1111/ele.12962. <http://doi.wiley.com/10.1111/ele.12962>.
- 27 Anderegg, W. R. L., and Coauthors, 2018b: Hydraulic diversity of forests regulates ecosystem resilience
28 during drought. *Nature*, **561**, 538–, doi:10.1038/s41586-018-0539-7.
- 29 Anderson-Teixeira, K. J., P. K. Snyder, T. E. Twine, S. V. Cuadra, M. H. Costa, and E. H. Delucia, 2012:
30 Climate-regulation services of natural and agricultural ecoregions of the Americas. *Nat. Clim. Chang.*,
31 **2**, 177–181, doi:10.1038/nclimate1346. <http://dx.doi.org/10.1038/nclimate1346>.
- 32 Anderson, R. G., and Coauthors, 2011: Biophysical considerations in forestry for climate protection. *Front.*
33 *Ecol. Environ.*, **9**, 174–182, doi:10.1890/090179.
- 34 Andreae, and Coauthors, 2002: Biogeochemical cycling of carbon, water, energy, trace gases, and aerosols in
35 Amazonia: The LBA-EUSTACH experiments. *J. Geophys. Res. D Atmos.*, **107**, 1998–2000,
36 doi:10.1029/2001JD000524.
- 37 Andreae, M. O., and Coauthors, 2018: Aerosol characteristics and particle production in the upper
38 troposphere over the Amazon Basin. *Atmos. Chem. Phys.*, **18**, 921–961, doi:10.5194/acp-18-921-2018.
- 39 Angel, S., J. Parent, D. L. Civco, A. Blei, and D. Potere, 2011: The dimensions of global urban expansion:
40 Estimates and projections for all countries, 2000–2050. *Prog. Plann.*, **75**, 53–107,
41 doi:<https://doi.org/10.1016/j.progress.2011.04.001>.
42 <http://www.sciencedirect.com/science/article/pii/S0305900611000109>.
- 43 Ansmann, A., and Coauthors, 2018: Extreme levels of Canadian wildfire smoke in the stratosphere over
44 central Europe on 21–22 August~2017. *Atmos. Chem. Phys.*, **18**, 11831–11845, doi:10.5194/acp-18-
45 11831-2018.
- 46 Aragão, L. E. O. C., and Coauthors, 2018: 21st Century drought-related fires counteract the decline of
47 Amazon deforestation carbon emissions. *Nat. Commun.*, **9**, 536, doi:10.1038/s41467-017-02771-y.
- 48 Archibald, S., C. E. R. Lehmann, J. L. Gómez-Dans, and R. A. Bradstock, 2013: Defining pyromes and
49 global syndromes of fire regimes. *Proc. Natl. Acad. Sci.*, **110**, 6442–6447,
50 doi:10.1073/pnas.1211466110. <https://www.pnas.org/content/110/16/6442>.
- 51 Argüeso, D., J. P. Evans, L. Fita, and K. J. Bormann, 2014: Temperature response to future urbanization and
52 climate change. *Clim. Dyn.*, **42**, 2183–2199, doi:10.1007/s00382-013-1789-6.
- 53 Arnell, N. W., and S. N. Gosling, 2016: The impacts of climate change on river flood risk at the global scale.
54 *Clim. Change*, **134**, 387–401, doi:10.1007/s10584-014-1084-5.
55 <http://link.springer.com/10.1007/s10584-014-1084-5> (Accessed October 31, 2018).
- 56 Arneth, A., and Coauthors, 2010: Terrestrial biogeochemical feedbacks in the climate system. *Nat. Geosci.*,

- 1 **3**, 525–532, doi:10.1038/ngeo905.
- 2 Arneeth, A., and Coauthors, 2017a: Historical carbon dioxide emissions caused by land-use changes are
3 possibly larger than assumed. *Nat. Geosci.*, **10**, 79. <http://dx.doi.org/10.1038/ngeo2882>.
- 4 Arneeth, A., and Coauthors, 2017b: Historical carbon dioxide emissions caused by land-use changes are
5 possibly larger than assumed. *Nat. Geosci.*, **10**, 79–84, doi:10.1038/ngeo2882.
6 <http://dx.doi.org/10.1038/ngeo2882>.
- 7 Arora, V. K., and G. J. Boer, 2010: Uncertainties in the 20th century carbon budget associated with land use
8 change. *Glob. Chang. Biol.*, doi:10.1111/j.1365-2486.2010.02202.x.
- 9 Arora, V. K., and A. Montenegro, 2011: Small temperature benefits provided by realistic afforestation
10 efforts. *Nat. Geosci.*, **4**, 514–518, doi:10.1038/ngeo1182.
- 11 ———, and J. R. Melton, 2018: Reduction in global area burned and wildfire emissions since 1930s enhances
12 carbon uptake by land. *Nat. Commun.*, doi:10.1038/s41467-018-03838-0.
- 13 Ariso, B. K., G. Mengistu Tsidu, G. H. Stoffberg, and T. Tadesse, 2018: Influence of urbanization-driven
14 land use/cover change on climate: The case of Addis Ababa, Ethiopia. *Phys. Chem. Earth*,
15 doi:10.1016/j.pce.2018.02.009.
- 16 Artaxo, P., and Coauthors, 2013: Atmospheric aerosols in Amazonia and land use change: From natural
17 biogenic to biomass burning conditions. *Faraday Discuss.*, **165**, 203–235, doi:10.1039/c3fd00052d.
- 18 Asefaw, B. A., H. J. W., T. M. S., H. John, A. A. Berhe, J. W. Harden, M. S. Torn, and J. Harte, 2008:
19 Linking soil organic matter dynamics and erosion-induced terrestrial carbon sequestration at different
20 landform positions. *J. Geophys. Res. Biogeosciences*, **113**, 1–12, doi:10.1029/2008JG000751.
21 <https://doi.org/10.1029/2008JG000751>.
- 22 Ashworth, K., G. Folberth, C. N. Hewitt, and O. Wild, 2012: Impacts of near-future cultivation of biofuel
23 feedstocks on atmospheric composition and local air quality. *Atmos. Chem. Phys.*, **12**, 919–939,
24 doi:10.5194/acp-12-919-2012. <http://www.atmos-chem-phys.net/12/919/2012/> (Accessed April 11,
25 2012).
- 26 Asseng, S., and Coauthors, 2015: Rising temperatures reduce global wheat production. *Nat. Clim. Chang.*, **5**,
27 143–147, doi:10.1038/nclimate2470.
- 28 Atkin, O. K., and Coauthors, 2015: Global variability in leaf respiration in relation to climate, plant
29 functional types and leaf traits. *New Phytol.*, **206**, 614–636, doi:10.1111/nph.13253.
- 30 Auffret, M. D., and Coauthors, 2016: The Role of Microbial Community Composition in Controlling Soil
31 Respiration Responses to Temperature. *PLoS One*, **11**, e0165448, doi:10.1371/journal.pone.0165448.
32 <http://dx.plos.org/10.1371/journal.pone.0165448>.
- 33 Augustin, J., J. Couwenberg, and M. Minke, 2011: Peatlands and greenhouse gases. *Carbon credits from*
34 *peatland rewetting: climate - biodiversity - land use*, T. Parish, F., Sirin, A., Charman, D., Joosten, H.,
35 Minayeva and L. Silvius, M. and Stringer, Eds., Global Environment Centre and Wetlands
36 International, Kuala Lumpur and Wageningen, 13–19
37 http://www.imcg.net/media/download_gallery/books/assessment_peatland.pdf.
- 38 Averill, C., B. L. Turner, and A. C. Finzi, 2014: Mycorrhiza-mediated competition between plants and
39 decomposers drives soil carbon storage. *Nature*, **505**, 543–545, doi:10.1038/nature12901.
40 <http://dx.doi.org/10.1038/nature12901>.
- 41 Avila, F. B., A. J. Pitman, M. G. Donat, L. V. Alexander, and G. Abramowitz, 2012: Climate model
42 simulated changes in temperature extremes due to land cover change. *J. Geophys. Res. Atmos.*, **117**, 1–
43 19, doi:10.1029/2011JD016382.
- 44 Avissar, R., and D. Werth, 2005: Global Hydroclimatological Teleconnections Resulting from Tropical
45 Deforestation. *J. Hydrometeorol.*, **6**, 134–145, doi:10.1175/JHM406.1.
- 46 Avitabile, V., M. Herold, G. B. M. Heuvelink, A. G. P. de B. Jong, and V. G. Laurin, 2016: An integrated
47 pan-tropical biomass map using multiple reference datasets. *Glob. Chang. Biol.*, **22**, 1406–1420.
- 48 Babst, F., M. Carrer, B. Poulter, C. Urbinati, B. Neuwirth, and D. Frank, 2012: 500 years of regional forest
49 growth variability and links to climatic extreme events in Europe. *Environ. Res. Lett.*, **7**, 45705,
50 doi:10.1088/1748-9326/7/4/045705. [http://stacks.iop.org/1748-
51 9326/7/i=4/a=045705?key=crossref.0c9cfeb416c9dbc92fd8de38d392d4de](http://stacks.iop.org/1748-9326/7/i=4/a=045705?key=crossref.0c9cfeb416c9dbc92fd8de38d392d4de) (Accessed October 30,
52 2018).
- 53 Baccini, A., and Coauthors, 2012: Estimated carbon dioxide emissions from tropical deforestation improved
54 by carbon-density maps. *Nat. Clim. Chang.*, **2**, doi:10.1038/nclimate1354.
- 55 ———, W. Walker, L. Carvalho, M. Farina, D. Sulla-Menashe, and R. A. Houghton, 2017: Tropical forests are
56 a net carbon source based on aboveground measurements of gain and loss. *Science (80-.)*, **358**, 230–

- 1 234, doi:10.1126/science.aam5962.
- 2 Bader, D. A., and Coauthors, 2018: Urban Climate Science. *Climate Change and Cities: Second Assessment*
- 3 *Report of the Urban Climate Change Research Network*, C. Rosenzweig, W. Solecki, P. Romer-
- 4 Lankao, S. Mehotra, S. Dhakal, and S. Ali Ibrahim, Eds., Cambridge University Press, Cambridge, 27–
- 5 60.
- 6 Bador, M., L. Terray, and J. Boé, 2016: Emergence of human influence on summer record-breaking
- 7 temperatures over Europe. *Geophys. Res. Lett.*, **43**, 404–412, doi:10.1002/2015GL066560.
- 8 <http://doi.wiley.com/10.1002/2015GL066560> (Accessed October 28, 2018).
- 9 Bae, J., and Y. Ryu, 2015: Land use and land cover changes explain spatial and temporal variations of the
- 10 soil organic carbon stocks in a constructed urban park. *Landsc. Urban Plan.*, **136**, 57–67,
- 11 doi:10.1016/j.landurbplan.2014.11.015. [http://www.scopus.com/inward/record.url?eid=2-s2.0-](http://www.scopus.com/inward/record.url?eid=2-s2.0-84919904060&partnerID=40&md5=cf5f6eed614773879b2092e47312e484)
- 12 [84919904060&partnerID=40&md5=cf5f6eed614773879b2092e47312e484](http://www.scopus.com/inward/record.url?eid=2-s2.0-84919904060&partnerID=40&md5=cf5f6eed614773879b2092e47312e484).
- 13 Bagley, J. E., L. M. Kueppers, D. P. Billesbach, I. N. Williams, S. C. Biraud, and M. S. Torn, 2017: The
- 14 influence of land cover on surface energy partitioning and evaporative fraction regimes in the U.S.
- 15 Southern Great Plains. *J. Geophys. Res. Atmos.*, **122**, 5793–5807, doi:10.1002/2017JD026740.
- 16 <http://doi.wiley.com/10.1002/2017JD026740>.
- 17 Bajželj, B., and K. Richards, 2014: The Positive Feedback Loop between the Impacts of Climate Change and
- 18 Agricultural Expansion and Relocation. *Land*, **3**, 898–916, doi:10.3390/land3030898.
- 19 <http://www.mdpi.com/2073-445X/3/3/898/>.
- 20 Baker, J. S., C. M. Wade, B. L. Sohngen, S. Ohrel, and A. A. Fawcett, 2019: Potential complementarity
- 21 between forest carbon sequestration incentives and biomass energy expansion. *Energy Policy*,
- 22 doi:10.1016/j.enpol.2018.10.009.
- 23 Bala, G., K. Caldeira, M. Wickett, T. J. Phillips, D. B. Lobell, C. Delire, and A. Mirin, 2007: Combined
- 24 climate and carbon-cycle effects of large-scale deforestation. *Proc. Natl. Acad. Sci.*, **104**, 6550–6555,
- 25 doi:10.1073/pnas.0608998104.
- 26 Ballantyne, A., and Coauthors, 2017: Accelerating net terrestrial carbon uptake during the warming hiatus
- 27 due to reduced respiration. *Nat. Clim. Chang.*, **7**, 148–152, doi:10.1038/nclimate3204.
- 28 <http://dx.doi.org/10.1038/nclimate3204>.
- 29 Ban, N., J. Schmidli, and C. Schär, 2015: Heavy precipitation in a changing climate: Does short-term
- 30 summer precipitation increase faster? *Geophys. Res. Lett.*, **42**, 1165–1172,
- 31 doi:10.1002/2014GL062588.
- 32 [https://agupubs.onlinelibrary.wiley.com/doi/full/10.1002/2014GL062588%4010.1002/%28ISSN%2919](https://agupubs.onlinelibrary.wiley.com/doi/full/10.1002/2014GL062588%4010.1002/%28ISSN%291944-8007.2015EdHighlights)
- 33 [44-8007.2015EdHighlights](https://agupubs.onlinelibrary.wiley.com/doi/full/10.1002/2014GL062588%4010.1002/%28ISSN%291944-8007.2015EdHighlights) (Accessed April 11, 2019).
- 34 Barbero-Sierra, C., M. J. Marques, and M. Ruíz-Pérez, 2013: The case of urban sprawl in Spain as an active
- 35 and irreversible driving force for desertification. *J. Arid Environ.*, **90**, 95–102,
- 36 doi:<https://doi.org/10.1016/j.jaridenv.2012.10.014>.
- 37 <http://www.sciencedirect.com/science/article/pii/S0140196312002820>.
- 38 Barnes, C. J., C. J. van der Gast, N. P. Mcnamara, R. Rowe, and G. D. Bending, 2018: Extreme rainfall
- 39 affects assembly of the root-associated fungal community. *New Phytol.*, doi:10.1111/nph.14990.
- 40 <http://doi.wiley.com/10.1111/nph.14990> (Accessed May 25, 2018).
- 41 Baron, R. E., W. D. Montgomery, and S. D. Tuladhar, 2009: *An Analysis of Black Carbon Mitigation as a*
- 42 *Response to Climate Change a Response to Climate Change*. 29pp pp.
- 43 Basche, A. D., F. E. Miguez, T. C. Kaspar, and M. J. Castellano, 2014: Do cover crops increase or decrease
- 44 nitrous oxide emissions? A meta-analysis. *J. Soil Water Conserv.*, **69**, 471–482,
- 45 doi:10.2489/jswc.69.6.471.
- 46 Baskaran, P., R. Hyvönen, S. L. Berglund, K. E. Clemmensen, G. I. Jørgensen, B. D. Lindahl, and S.
- 47 Manzoni, 2017: Modelling the influence of ectomycorrhizal decomposition on plant nutrition and soil
- 48 carbon sequestration in boreal forest ecosystems. *New Phytol.*, **213**, 1452–1465,
- 49 doi:10.1111/nph.14213. <https://doi.org/10.1111/nph.14213>.
- 50 Bassu, S., and Coauthors, 2014: How do various maize crop models vary in their responses to climate
- 51 change factors? *Glob. Chang. Biol.*, **20**, 2301–2320, doi:10.1111/gcb.12520.
- 52 Bathiany, S., M. Claussen, V. Brovkin, T. Raddatz, and V. Gayler, 2010a: Combined biogeophysical and
- 53 biogeochemical effects of large-scale forest cover changes in the MPI earth system model.
- 54 *Biogeosciences*, **7**, 1383–1399, doi:10.5194/bg-7-1383-2010.
- 55 —, —, —, —, and —, 2010b: Combined biogeophysical and biogeochemical effects of large-
- 56 scale forest cover changes in the MPI earth system model. *Biogeosciences*, **7**, 1383–1399,

- 1 doi:10.5194/bg-7-1383-2010.
- 2 Battipaglia, G., M. Saurer, P. Cherubini, C. Calfapietra, H. R. McCarthy, R. J. Norby, and M. Francesca
3 Cotrufo, 2013: Elevated CO₂ increases tree-level intrinsic water use efficiency: Insights from carbon
4 and oxygen isotope analyses in tree rings across three forest FACE sites. *New Phytol.*, **197**, 544–554,
5 doi:10.1111/nph.12044.
- 6 Beach, R. H., J. Creason, S. B. Ohrel, S. Ragnauth, S. Ogle, C. Li, P. Ingraham, and W. Salas, 2015: Global
7 mitigation potential and costs of reducing agricultural non-CO₂ greenhouse gas emissions through
8 2030. *J. Integr. Environ. Sci.*, **12**, 87–105, doi:10.1080/1943815X.2015.1110183.
- 9 Beer, C., and Coauthors, 2010: Terrestrial gross carbon dioxide uptake: global distribution and covariation
10 with climate. *Science*, **329**, 834–838, doi:10.1126/science.1184984.
11 <http://www.ncbi.nlm.nih.gov/pubmed/20603496> (Accessed November 2, 2018).
- 12 Beer, T., 2018: The Impact of Extreme Weather Events on Food Security. Springer, Cham, 121–133
13 http://link.springer.com/10.1007/978-3-319-56469-2_8 (Accessed May 25, 2018).
- 14 Beerling, D. J., and Coauthors, Farming with crops and rocks to address global climate, food and soil
15 security. *nature.com*,.
- 16 Beerling, D. J., and Coauthors, 2018: Farming with crops and rocks to address global climate, food and soil
17 security. *Nat. Plants*, **4**, 138–147, doi:10.1038/s41477-018-0108-y.
- 18 Bento, A. M., and R. Klotz, 2014: Climate policy decisions require policy-based lifecycle analysis. *Environ.*
19 *Sci. Technol.*, doi:10.1021/es405164g.
- 20 Berdanier, A. B., and J. S. Clark, 2016: Multiyear drought-induced morbidity preceding tree death in
21 southeastern U.S. forests. *Ecol. Appl.*, **26**, 17–23.
- 22 Berg, A., and J. Sheffield, 2018: Climate Change and Drought: the Soil Moisture Perspective. *Curr. Clim.*
23 *Chang. Reports*, **4**, 180–191, doi:10.1007/s40641-018-0095-0.
24 <http://link.springer.com/10.1007/s40641-018-0095-0> (Accessed October 28, 2018).
- 25 —, and Coauthors, 2015: Interannual coupling between summertime surface temperature and precipitation
26 over land: Processes and implications for climate change. *J. Clim.*, **28**, 1308–1328, doi:10.1175/JCLI-
27 D-14-00324.1. <http://journals.ametsoc.org/doi/10.1175/JCLI-D-14-00324.1> (Accessed November 1,
28 2018).
- 29 —, and Coauthors, 2016: Land-atmosphere feedbacks amplify aridity increase over land under global
30 warming. *Nat. Clim. Chang.*, **6**, 869–874, doi:10.1038/nclimate3029.
- 31 Berg, E. E., J. David Henry, C. L. Fastie, A. D. De Volder, and S. M. Matsuoka, 2006: Spruce beetle
32 outbreaks on the Kenai Peninsula, Alaska, and Kluane National Park and Reserve, Yukon Territory:
33 Relationship to summer temperatures and regional differences in disturbance regimes. *For. Ecol.*
34 *Manage.*, **227**, 219–232, doi:10.1016/j.foreco.2006.02.038.
35 <https://www.sciencedirect.com/science/article/pii/S0378112706001599> (Accessed April 2, 2019).
- 36 Berg, P., C. Moseley, and J. O. Haerter, 2013: Strong increase in convective precipitation in response to
37 higher temperatures. *Nat. Geosci.*, **6**, 181–185, doi:10.1038/ngeo1731.
- 38 Bergamaschi, P., and Coauthors, 2013: Atmospheric CH₄ in the first decade of the 21st century:
39 Inverse modeling analysis using SCIAMACHY satellite retrievals and NOAA surface measurements. *J.*
40 *Geophys. Res. Atmos.*, **118**, 7350–7369, doi:10.1002/jgrd.50480.
- 41 Berhe, A. A., J. Harte, J. W. Harden, and M. S. Torn, 2007: The Significance of the Erosion-induced
42 Terrestrial Carbon Sink. *Bioscience*, **57**, 337–346, doi:10.1641/B570408.
43 [http://academic.oup.com/bioscience/article/57/4/337/220522/The-Significance-of-the-Erosion-induced-
44 Terrestrial](http://academic.oup.com/bioscience/article/57/4/337/220522/The-Significance-of-the-Erosion-induced-Terrestrial).
- 45 Berndes, G., S. Ahlgren, P. Börjesson, and A. L. Cowie, 2013: Bioenergy and land use change—state of the
46 art. *Wiley Interdiscip. Rev. Energy Environ.*, **2**, 282–303, doi:10.1002/wene.41.
47 <http://dx.doi.org/10.1002/wene.41>.
- 48 Bernier, P., and D. Paré, 2013: Using ecosystem CO₂ measurements to estimate the timing and magnitude of
49 greenhouse gas mitigation potential of forest bioenergy. *GCB Bioenergy*, **5**, 67–72, doi:10.1111/j.1757-
50 1707.2012.01197.x.
- 51 Berry, J. A., D. J. Beerling, and P. J. Franks, 2010: Stomata: Key players in the earth system, past and
52 present. *Curr. Opin. Plant Biol.*, **13**, 233–240, doi:10.1016/j.pbi.2010.04.013.
- 53 Bertram, C., and Coauthors, 2018: Targeted policies can compensate most of the increased sustainability
54 risks in 1.5 °C mitigation scenarios. *Environ. Res. Lett.*, **13**, 64038, doi:10.1088/1748-9326/aac3ec.
55 [http://stacks.iop.org/1748-9326/13/i=6/a=064038?key=crossref.ff7b320414e7bb8c685ea4743f5d4c28
56](http://stacks.iop.org/1748-9326/13/i=6/a=064038?key=crossref.ff7b320414e7bb8c685ea4743f5d4c28) (Accessed October 15, 2018).

- 1 Betts, R. a., P. D. Falloon, K. K. Goldewijk, and N. Ramankutty, 2007: Biogeophysical effects of land use on
2 climate: Model simulations of radiative forcing and large-scale temperature change. *Agric. For.*
3 *Meteorol.*, **142**, 216–233, doi:10.1016/j.agrformet.2006.08.021.
- 4 Betts, R. A., 2001: Biogeophysical impacts of land use on present-day climate: Near-surface temperature
5 change and radiative forcing. *Atmos. Sci. Lett.*, **2**, 1–13, doi:10.1006/asle.2001.0023.
- 6 Di Biagio, C., and Coauthors, 2017: Global scale variability of the mineral dust long-wave refractive index: a
7 new dataset of in situ measurements for climate modeling and remote sensing. *Atmos. Chem. Phys.*, **17**,
8 1901–1929, doi:10.5194/acp-17-1901-2017.
- 9 Birdsey, R., and Y. Pan, 2015: *Trends in management of the world's forests and impacts on carbon stocks*.
- 10 Blanco-Canqui, H., and R. Lal, 2004: Mechanisms of carbon sequestration in soil aggregates. *CRC. Crit.*
11 *Rev. Plant Sci.*, **23**, 481–504, doi:10.1080/07352680490886842.
12 <https://doi.org/10.1080/07352680490886842>.
- 13 Boettle, M., D. Rybski, and J. P. Kropp, 2016: Quantifying the effect of sea level rise and flood defence
14 – A point process perspective on coastal flood damage. *Nat. Hazards Earth Syst. Sci.*, **16**, 559–
15 576, doi:10.5194/nhess-16-559-2016.
- 16 Bohrer, G., H. Mourad, T. A. Laursen, D. Drewry, R. Avissar, D. Poggi, R. Oren, and G. G. Katul, 2005:
17 Finite element tree crown hydrodynamics model (FETCH) using porous media flow within branching
18 elements: A new representation of tree hydrodynamics. *Water Resour. Res.*, **41**,
19 doi:10.1029/2005wr004181.
- 20 Boisier, J. P., and Coauthors, 2012: Attributing the impacts of land-cover changes in temperate regions on
21 surface temperature and heat fluxes to specific causes: Results from the first LUCID set of simulations.
22 *J. Geophys. Res. Atmos.*, **117**, 1–16, doi:10.1029/2011JD017106.
- 23 Boisvert-Marsh, L., C. Périé, and S. de Blois, 2014: Shifting with climate? Evidence for recent changes in
24 tree species distribution at high latitudes. *Ecosphere*, **5**, art83, doi:10.1890/ES14-00111.1.
25 <http://doi.wiley.com/10.1890/ES14-00111.1> (Accessed March 12, 2019).
- 26 Bonan, G. B., 2008a: Forests and climate change: Forcings, feedbacks, and the climate benefits of forests.
27 *Science (80-.)*, **320**, 1444–1449, doi:10.1126/science.1155121.
- 28 Bonan, G. B., 2008b: Forests and Climate Change: Forcings, Feedbacks, and the Climate Benefits of Forests.
29 *Science (80-.)*, **320**, 1444 LP-1449. <http://science.sciencemag.org/content/320/5882/1444.abstract>.
- 30 Bond-Lamberty, B., and A. Thomson, 2010: Temperature-associated increases in the global soil respiration
31 record. *Nature*, **464**, 579–582, doi:10.1038/nature08930. <http://www.nature.com/articles/nature08930>
32 (Accessed November 2, 2018).
- 33 —, V. L. Bailey, M. Chen, C. M. Gough, and R. Vargas, 2018: Globally rising soil heterotrophic
34 respiration over recent decades. *Nature*, **560**, 80–83, doi:10.1038/s41586-018-0358-x.
35 <http://dx.doi.org/10.1038/s41586-018-0358-x>.
- 36 Bond, T. C., and Coauthors, 2013: Bounding the role of black carbon in the climate system: A scientific
37 assessment. *J. Geophys. Res. Atmos.*, **118**, 5380–5552, doi:10.1002/jgrd.50171.
38 <https://doi.org/10.1002/jgrd.50171>.
- 39 Bond, W. J., F. I. Woodward, G. F. Midgley, B. W. J., W. F. I., and M. G. F., 2004: The global distribution
40 of ecosystems in a world without fire. *New Phytol.*, **165**, 525–538, doi:10.1111/j.1469-
41 8137.2004.01252.x.
- 42 Bonfils, C., and D. Lobell, 2007: Empirical evidence for a recent slowdown in irrigation-induced cooling.
43 *Proc. Natl. Acad. Sci.*, **104**, 13582–13587, doi:10.1073/pnas.0700144104.
44 <http://www.pnas.org/cgi/doi/10.1073/pnas.0700144104>.
- 45 Boone, A. A., and Coauthors, 2016: The regional impact of Land-Use Land-cover Change (LULCC) over
46 West Africa from an ensemble of global climate models under the auspices of the WAMME2 project.
47 *Clim. Dyn.*, **47**, 3547–3573, doi:10.1007/s00382-016-3252-y.
- 48 Borchard, N., and Coauthors, 2019: Biochar, soil and land-use interactions that reduce nitrate leaching and
49 N2O emissions: A meta-analysis. *Sci. Total Environ.*, doi:10.1016/j.scitotenv.2018.10.060.
- 50 Borodina, A., E. M. Fischer, and R. Knutti, 2017: Models are likely to underestimate increase in heavy
51 rainfall in the extratropical regions with high rainfall intensity. *Geophys. Res. Lett.*, **44**, 7401–7409,
52 doi:10.1002/2017GL074530. <http://doi.wiley.com/10.1002/2017GL074530> (Accessed May 16, 2018).
- 53 Bosello, F., C. Carraro, and E. De Cian, 2009: *An Analysis of Adaptation as a Response to Climate Change*.
54 29pp pp.
- 55 Bossio, D. A., and K. M. Scow, 1998: Impacts of carbon and flooding on soil microbial communities:
56 Phospholipid fatty acid profiles and substrate utilization patterns. *Microb. Ecol.*, **35**, 265–278,

- 1 doi:10.1007/s002489900082. <http://link.springer.com/10.1007/s002489900082> (Accessed May 25,
2 2018).
- 3 Boucher, O., and Coauthors, 2013: Clouds and aerosols. *Climate Change 2013 the Physical Science Basis:
4 Working Group I Contribution to the Fifth Assessment Report of the Intergovernmental Panel on
5 Climate Change*, T.F. Stocker et al., Eds., Vol. 9781107057 of, Cambridge University Press,
6 Cambridge, UK, 571–658.
- 7 Boucher, O., and Coauthors, 2016: Jury is still out on the radiative forcing by black carbon. *Proc. Natl.
8 Acad. Sci.*, **113**, E5092–E5093, doi:10.1073/pnas.1607005113.
9 <http://www.pnas.org/lookup/doi/10.1073/pnas.1607005113>.
- 10 Bourtsoukidis, E., and Coauthors, 2018: Strong sesquiterpene emissions from Amazonian soils. *Nat.
11 Commun.*, **9**, doi:10.1038/s41467-018-04658-y.
- 12 Bousquet, P., and Coauthors, 2006: Contribution of anthropogenic and natural sources to atmospheric
13 methane variability. *Nature*, **443**, 439–443, doi:10.1038/nature05132.
- 14 —, and Coauthors, 2011a: Source attribution of the changes in atmospheric methane for 2006–2008.
15 *Atmos. Chem. Phys.*, **11**, 3689–3700, doi:10.5194/acp-11-3689-2011.
- 16 Bousquet, P., and Coauthors, 2011b: and Physics. 3689–3700, doi:10.5194/acp-11-3689-2011.
- 17 Bouwman, A. F., and Coauthors, 2013: Global trends and uncertainties in terrestrial denitrification and N₂O
18 emissions. *Philos. Trans. R. Soc. B Biol. Sci.*, **368**, 20130112–20130112, doi:10.1098/rstb.2013.0112.
19 <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2013.0112>.
- 20 Bowman, D. M. J. S., and Coauthors, 2011a: The human dimension of fire regimes on Earth. *J. Biogeogr.*,
21 **38**, 2223–2236, doi:10.1111/j.1365-2699.2011.02595.x. [http://doi.wiley.com/10.1111/j.1365-
22 2699.2011.02595.x](http://doi.wiley.com/10.1111/j.1365-2699.2011.02595.x).
- 23 —, and Coauthors, 2011b: The human dimension of fire regimes on Earth. *J. Biogeogr.*, **38**, 2223–2236,
24 doi:10.1111/j.1365-2699.2011.02595.x.
- 25 Bowman, D. M. J. S., G. J. Williamson, J. T. Abatzoglou, C. A. Kolden, M. A. Cochrane, and A. M. S.
26 Smith, 2017: Human exposure and sensitivity to globally extreme wildfire events. *Nat. Ecol. & Evol.*, **1**, 58.
27
- 28 Boysen, L. R., V. Brovkin, V. K. Arora, P. Cadule, N. de Noblet-Ducoudré, E. Kato, J. Pongratz, and V.
29 Gayler, 2014: Global and regional effects of land-use change on climate in 21st century simulations
30 with interactive carbon cycle. *Earth Syst. Dyn. Discuss.*, **5**, 443–472, doi:10.5194/esdd-5-443-2014.
- 31 Boysen, L. R., W. Lucht, and D. Gerten, 2017: Trade-offs for food production, nature conservation and
32 climate limit the terrestrial carbon dioxide removal potential. *Glob. Chang. Biol.*, **23**, 4303–4317,
33 doi:10.1111/gcb.13745.
- 34 Bozzi, E., L. Genesio, P. Toscano, M. Pieri, and F. Miglietta, 2015: Mimicking biochar-albedo feedback in
35 complex Mediterranean agricultural landscapes. *Environ. Res. Lett.*, **10**, doi:10.1088/1748-
36 9326/10/8/084014.
- 37 Bracho-Nunez, A., S. Welter, M. Staudt, and J. Kesselmeier, 2011: Plant-specific volatile organic compound
38 emission rates from young and mature leaves of Mediterranean vegetation. *J. Geophys. Res. Atmos.*,
39 **116**, 1–13, doi:10.1029/2010JD015521.
- 40 Braconnot, P., S. P. Harrison, M. Kageyama, P. J. Bartlein, V. Masson-Delmotte, A. Abe-Ouchi, B. Otto-
41 Bliessner, and Y. Zhao, 2012: Evaluation of climate models using palaeoclimatic data. *Nat. Clim.
42 Chang.*, **2**, 417–424, doi:10.1038/nclimate1456.
- 43 Bradford, M. A., and Coauthors, 2008: Thermal adaptation of soil microbial respiration to elevated
44 temperature. *Ecol. Lett.*, **11**, 1316–1327, doi:10.1111/j.1461-0248.2008.01251.x.
- 45 Brando, P. M., D. C. Nepstad, E. A. Davidson, S. E. Trumbore, D. Ray, and P. Camargo, 2008: Drought
46 effects on litterfall, wood production and belowground carbon cycling in an Amazon forest: results of a
47 throughfall reduction experiment. *Philos. Trans. R. Soc. B Biol. Sci.*, **363**, 1839–1848,
48 doi:10.1098/rstb.2007.0031. <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2007.0031>
49 (Accessed May 21, 2018).
- 50 Brando, P. M., and Coauthors, 2014: Abrupt increases in Amazonian tree mortality due to drought-fire
51 interactions. *Proc. Natl. Acad. Sci.*, **111**, 6347–6352, doi:10.1073/pnas.1305499111.
- 52 Braun, M., and Coauthors, 2016: A holistic assessment of greenhouse gas dynamics from forests to the
53 effects of wood products use in Austria. *Carbon Manag.*, **7**, 271–283,
54 doi:10.1080/17583004.2016.1230990.
- 55 Bremer, D. J., and J. M. Ham, 1999: Effect of spring burning on the surface energy balance in a tallgrass
56 prairie. *Agric. For. Meteorol.*, **97**, 43–54, doi:10.1016/S0168-1923(99)00034-9.

- 1 <http://www.sciencedirect.com/science/article/pii/S0168192399000349>.
- 2 Bren d'Amour, C., and Coauthors, 2016: Future urban land expansion and implications for global croplands.
3 *Proc. Natl. Acad. Sci.*, **114**, 8939–8944, doi:10.1073/pnas.1606036114.
- 4 Breshears, D. D., and Coauthors, 2005: Regional vegetation die-off in response to global-change-type
5 drought. *Proc. Natl. Acad. Sci.*, **102**, 15144–15148, doi:10.1073/pnas.0505734102.
6 <http://www.pnas.org/cgi/doi/10.1073/pnas.0505734102>.
- 7 Brienen, R. J. W., and Coauthors, 2015: Long-term decline of the Amazon carbon sink. *Nature*,
8 doi:10.1038/nature14283.
- 9 Bright, R. M., K. Zhao, R. B. Jackson, and F. Cherubini, 2015: Quantifying surface albedo and other direct
10 biogeophysical climate forcings of forestry activities. *Glob. Chang. Biol.*, **21**, 3246–3266,
11 doi:10.1111/gcb.12951.
- 12 ———, E. Davin, T. O'Halloran, J. Pongratz, K. Zhao, and A. Cescatti, 2017: Local temperature response to
13 land cover and management change driven by non-radiative processes. *Nat. Clim. Chang.*, **7**, 296–302,
14 doi:10.1038/nclimate3250.
- 15 Brodribb, T. J., 2009: Xylem hydraulic physiology: The functional backbone of terrestrial plant productivity.
16 *Plant Sci.*, **177**, 245–251, doi:10.1016/j.plantsci.2009.06.001.
17 <http://www.sciencedirect.com/science/article/pii/S0168945209001599>.
- 18 Brodt, S., K. J. Kramer, A. Kendall, and G. Feenstra, 2013: Comparing environmental impacts of regional
19 and national-scale food supply chains: A case study of processed tomatoes. *Food Policy*, **42**, 106–114,
20 doi:<https://doi.org/10.1016/j.foodpol.2013.07.004>.
21 <http://www.sciencedirect.com/science/article/pii/S0306919213001012>.
- 22 Brovkin, V., S. Sitch, W. von Bloh, M. Claussen, E. Bauer, and W. Cramer, 2004: Role of land cover
23 changes for atmospheric CO₂ increase and climate change during the last 150 years. *Glob. Chang.*
24 *Biol.*, **10**, 1253–1266, doi:10.1111/j.1365-2486.2004.00812.x.
- 25 ———, and Coauthors, 2006: Biogeophysical effects of historical land cover changes simulated by six Earth
26 system models of intermediate complexity. *Clim. Dyn.*, **26**, 587–600, doi:10.1007/s00382-005-0092-6.
- 27 Brovkin, V., and Coauthors, 2013a: Effect of anthropogenic land-use and land-cover changes on climate and
28 land carbon storage in CMIP5 projections for the twenty-first century. *J. Clim.*, **26**, 6859–6881,
29 doi:10.1175/JCLI-D-12-00623.1.
- 30 ———, and Coauthors, 2013b: Effect of anthropogenic land-use and land-cover changes on climate and land
31 carbon storage in CMIP5 projections for the twenty-first century. *J. Clim.*, **26**, 6859–6881,
32 doi:10.1175/JCLI-D-12-00623.1.
- 33 Brovkin, V., T. Pugh, E. Robertson, S. Bathiany, A. Arneth, and C. Jones, 2015: Cooling biogeophysical
34 effect of large-scale tropical deforestation in three Earth System models. *EGU General Assembly*
35 *Conference Abstracts*, Vol. 17 of *EGU General Assembly Conference Abstracts*, 8903.
- 36 Brown, M. E., and Coauthors, 2015: *Climate Change, Global Food Security, and the U.S. Food System*. 146
37 pp.
- 38 Brown, S., and R. J. Nicholls, 2015: Subsidence and human influences in mega deltas: The case of the
39 Ganges–Brahmaputra–Meghna. *Sci. Total Environ.*, **527–528**, 362–374,
40 doi:<https://doi.org/10.1016/j.scitotenv.2015.04.124>.
41 <http://www.sciencedirect.com/science/article/pii/S0048969715300589>.
- 42 Brutel-Vuilmet, C., M. Ménégoz, and G. Krinner, 2013: An analysis of present and future seasonal Northern
43 Hemisphere land snow cover simulated by CMIP5 coupled climate models. *Cryosphere*, **7**, 67–80,
44 doi:10.5194/tc-7-67-2013. <https://www.the-cryosphere.net/7/67/2013/>.
- 45 Bryn, A., and K. Potthoff, 2018: Elevational treeline and forest line dynamics in Norwegian mountain areas
46 – a review. *Landsc. Ecol.*, **33**, 1225–1245, doi:10.1007/s10980-018-0670-8.
47 <http://link.springer.com/10.1007/s10980-018-0670-8> (Accessed March 12, 2019).
- 48 Buchkowski, R. W., M. A. Bradford, A. S. Grandy, O. J. Schmitz, and W. R. Wieder, 2017: Applying
49 population and community ecology theory to advance understanding of belowground biogeochemistry.
50 *Ecol. Lett.*, **20**, 231–245, doi:10.1111/ele.12712. <https://doi.org/10.1111/ele.12712>.
- 51 Buitenhuis, E. T., P. Suntharalingam, and C. Le Quéré, 2017: Constraints on global oceanic emissions of
52 N₂O from observations and models. *Biogeosciences*
53 *Discuss.*, 1–23, doi:10.5194/bg-2017-193. <https://www.biogeosciences-discuss.net/bg-2017-193/>.
- 54 Buitenwerf, R., B. Sandel, S. Normand, A. Mimet, and J.-C. Svenning, 2018: Land surface greening suggests
55 vigorous woody regrowth throughout European semi-natural vegetation. *Glob. Chang. Biol.*, **24**, 5789–
56 5801, doi:10.1111/gcb.14451. <http://doi.wiley.com/10.1111/gcb.14451> (Accessed March 13, 2019).

- 1 Bullard, J. E., and Coauthors, 2016: High latitude dust in the Earth system *Reviews of Geophysics. Rev.*
2 *Geophys.*, **54**, 447–485, doi:10.1002/2016RG000518.Received.
- 3 Buotte, P. C., S. Levis, B. E. Law, T. W. Hudiburg, D. E. Rupp, and J. J. Kent, 2019: Near-future forest
4 vulnerability to drought and fire varies across the western United States. *Glob. Chang. Biol.*, **25**, 290–
5 303, doi:10.1111/gcb.14490.
- 6 Burakowski, E., A. Tawfik, A. Ouimette, L. Lepine, K. Novick, S. Ollinger, C. Zarzycki, and G. Bonan,
7 2018: The role of surface roughness, albedo, and Bowen ratio on ecosystem energy balance in the
8 Eastern United States. *Agric. For. Meteorol.*, **249**, 367–376, doi:10.1016/j.agrformet.2017.11.030.
- 9 Burkett, V. R., A. G. Suarez, M. Bindi, C. Conde, R. Mukerji, M. J. Prather, A. L. St. Clair, and G. W. Yohe,
10 2014: Point of departure. *Climate Change 2014: Impacts, Adaptation, and Vulnerability. Part A:
11 Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment Report of the
12 Inter- governmental Panel on Climate Change Aspects*, C.B. Field et al., Eds., Cambridge University
13 Press, Cambridge, United Kingdom and New York, NY, USA, 169–194.
- 14 Busch, F. A., and R. F. Sage, 2017: The sensitivity of photosynthesis to O₂ and CO₂ concentration identifies
15 strong Rubisco control above the thermal optimum. *New Phytol.*, **213**, 1036–1051,
16 doi:10.1111/nph.14258.
- 17 Busch, J., and J. Engelmann, 2017: Cost-effectiveness of reducing emissions from tropical deforestation,
18 2016–2050. *Environ. Res. Lett.*, **13**, 15001, doi:10.1088/1748-9326/aa907c. [http://stacks.iop.org/1748-
19 9326/13/i=1/a=015001?key=crossref.680f60c21ec50a4c4bfc7897153e6754](http://stacks.iop.org/1748-9326/13/i=1/a=015001?key=crossref.680f60c21ec50a4c4bfc7897153e6754).
- 20 Butt, N., P. A. De Oliveira, and M. H. Costa, 2011: Evidence that deforestation affects the onset of the rainy
21 season in Rondonia, Brazil. *J. Geophys. Res. Atmos.*, **116**, 2–9, doi:10.1029/2010JD015174.
- 22 Von Buttlar, J., and Coauthors, 2018: Impacts of droughts and extreme-temperature events on gross primary
23 production and ecosystem respiration: A systematic assessment across ecosystems and climate zones.
24 *Biogeosciences*, **15**, 1293–1318, doi:10.5194/bg-15-1293-2018.
25 <https://www.biogeosciences.net/15/1293/2018/> (Accessed October 16, 2018).
- 26 Byrne, M. P., and P. A. O’Gorman, 2013: Land-ocean warming contrast over a wide range of climates:
27 Convective quasi-equilibrium theory and idealized simulations. *J. Clim.*, **26**, 4000–4016,
28 doi:10.1175/JCLI-D-12-00262.1.
- 29 Byrne, M. P., and P. A. O’Gorman, 2015: The Response of Precipitation Minus Evapotranspiration to
30 Climate Warming: Why the “Wet-Get-Wetter, Dry-Get-Drier” Scaling Does Not Hold over Land. *J.
31 Clim.*, **28**, 8078–8092, doi:10.1175/JCLI-D-15-0369.1. <https://doi.org/10.1175/JCLI-D-15-0369.1>.
- 32 Cai, W., and Coauthors, 2014a: Increasing frequency of extreme El Niño events due to greenhouse warming.
33 *Nat. Clim. Chang.*, **4**, 111–116, doi:10.1038/nclimate2100.
- 34 —, A. Santoso, G. Wang, E. Weller, L. Wu, K. Ashok, Y. Masumoto, and T. Yamagata, 2014b: Increased
35 frequency of extreme Indian ocean dipole events due to greenhouse warming. *Nature*, **510**, 254–258,
36 doi:10.1038/nature13327. <http://www.nature.com/articles/nature13327> (Accessed November 1, 2018).
- 37 —, and Coauthors, 2015: Increased frequency of extreme La Nina events under greenhouse warming. *Nat.
38 Clim. Chang.*, **5**, 132–137, doi:10.1038/nclimate2492.
39 [http://www.nature.com/nclimate/journal/v5/n2/full/nclimate2492.html%5Cnhttp://www.nature.com/ncl
40 imate/journal/v5/n2/pdf/nclimate2492.pdf](http://www.nature.com/nclimate/journal/v5/n2/full/nclimate2492.html%5Cnhttp://www.nature.com/nclimate/journal/v5/n2/pdf/nclimate2492.pdf).
- 41 Cai, X., X. Zhang, and D. Wang, 2011: Land Availability for Biofuel Production. *Environ. Sci. Technol.*, **45**,
42 334–339, doi:10.1021/es103338e. <http://dx.doi.org/10.1021/es103338e>.
- 43 Calvin, K., M. Wise, P. Kyle, P. Patel, L. Clarke, and J. Edmonds, 2014: Trade-offs of different land and
44 bioenergy policies on the path to achieving climate targets. *Clim. Change*, **123**, 691–704,
45 doi:10.1007/s10584-013-0897-y. <http://link.springer.com/10.1007/s10584-013-0897-y> (Accessed
46 October 15, 2018).
- 47 —, and Coauthors, 2017: The SSP4: A world of deepening inequality. *Glob. Environ. Chang.*, **42**, 284–
48 296, doi:10.1016/J.GLOENVCHA.2016.06.010.
49 <https://www.sciencedirect.com/science/article/abs/pii/S095937801630084X> (Accessed November 4,
50 2018).
- 51 Calvo, A. I., C. Alves, A. Castro, V. Pont, A. M. Vicente, and R. Fraile, 2013: Research on aerosol sources
52 and chemical composition: Past, current and emerging issues. *Atmos. Res.*, **120–121**, 1–28,
53 doi:10.1016/j.atmosres.2012.09.021.
- 54 Camenzind, T., S. Hättenschwiler, K. K. Treseder, A. Lehmann, and M. C. Rillig, 2018: Nutrient limitation
55 of soil microbial processes in tropical forests. *Ecol. Monogr.*, **88**, 4–21, doi:10.1002/ecm.1279.
- 56 Camilloni, I., and M. Barrucand, 2012: Temporal variability of the Buenos Aires, Argentina, urban heat

- 1 island. *Theor. Appl. Climatol.*, **107**, 47–58, doi:10.1007/s00704-011-0459-z.
- 2 Campbell, J. E., D. B. Lobell, R. C. Genova, and C. B. Field, 2008: The Global Potential of Bioenergy on
3 Abandoned Agriculture Lands. *Environ. Sci. Technol.*, **42**, 5791–5794, doi:10.1021/es800052w.
- 4 Campbell, J. E., and Coauthors, 2017: Large historical growth in global terrestrial gross primary production.
5 *Nature*, **544**, 84–87, doi:10.1038/nature22030.
- 6 Campioli, M., and Coauthors, 2015: *Biomass production efficiency controlled by management in temperate
7 and boreal ecosystems*.
- 8 Campra, P., M. Garcia, Y. Canton, and A. Palacios-Orueta, 2008: Surface temperature cooling trends and
9 negative radiative forcing due to land use change toward greenhouse farming in southeastern Spain. *J.
10 Geophys. Res. Atmos.*, **113**, 1–10, doi:10.1029/2008JD009912.
- 11 Canadell, J. G., and Coauthors, 2007: Contributions to accelerating atmospheric CO₂ growth from economic
12 activity, carbon intensity, and efficiency of natural sinks. *Proc. Natl. Acad. Sci.*, **104**, 18866–18870,
13 doi:10.1073/pnas.0702737104. <http://www.pnas.org/cgi/doi/10.1073/pnas.0702737104>.
- 14 Cao, C., and Coauthors, 2016: Greening of the Earth and its drivers. *Nat. Clim. Chang.*, **6**, 791–795,
15 doi:10.1038/nclimate3004.
- 16 Cardenas, L. M., and Coauthors, 2019: Nitrogen use efficiency and nitrous oxide emissions from five UK
17 fertilised grasslands. *Sci. Total Environ.*, **661**, 696–710, doi:10.1016/j.scitotenv.2019.01.082.
18 <https://linkinghub.elsevier.com/retrieve/pii/S0048969719300890>.
- 19 Carey, J. C., and Coauthors, 2016: Temperature response of soil respiration largely unaltered with
20 experimental warming. *Proc. Natl. Acad. Sci.*, **113**, 13797–13802, doi:10.1073/pnas.1605365113.
21 <http://www.pnas.org/lookup/doi/10.1073/pnas.1605365113>.
- 22 Carissimo, B. C., A. H. Oort, and T. H. Vonder Haar, 1985: Estimating the Meridional Energy Transports in
23 the Atmosphere and Ocean. *J. Phys. Oceanogr.*, **15**, 82–91, doi:10.1175/1520-
24 0485(1985)015<0082:ETMETI>2.0.CO;2. [http://journals.ametsoc.org/doi/abs/10.1175/1520-
25 0485%281985%29015%3C0082%3AETMETI%3E2.0.CO%3B2](http://journals.ametsoc.org/doi/abs/10.1175/1520-0485%281985%29015%3C0082%3AETMETI%3E2.0.CO%3B2) (Accessed May 7, 2018).
- 26 Carlson, B. Z., M. C. Corona, C. Dentant, R. Bonet, W. Thuiller, and P. Choler, 2017: Observed long-term
27 greening of alpine vegetation - A case study in the French Alps. *Environ. Res. Lett.*, **12**, 114006,
28 doi:10.1088/1748-9326/aa84bd. [http://stacks.iop.org/1748-
29 9326/12/i=11/a=114006?key=crossref.c6497060c51d41c63d19f1d40ccad569](http://stacks.iop.org/1748-9326/12/i=11/a=114006?key=crossref.c6497060c51d41c63d19f1d40ccad569) (Accessed October 16,
30 2018).
- 31 Carrer, D., G. Pique, M. Ferlicoq, X. Ceamanos, and E. Ceschia, 2018: What is the potential of cropland
32 albedo management in the fight against global warming? A case study based on the use of cover crops.
33 *Environ. Res. Lett.*, **13**, 44030, doi:10.1088/1748-9326/aab650.
34 <http://iopscience.iop.org/article/10.1088/1748-9326/aab650>.
- 35 Carter, S., M. Herold, M. C. Rufino, K. Neumann, L. Kooistra, and L. Verchot, 2015: Mitigation of
36 agricultural emissions in the tropics: Comparing forest land-sparing options at the national level.
37 *Biogeosciences*, **12**, 4809–4825, doi:10.5194/bg-12-4809-2015.
38 <https://www.biogeosciences.net/12/4809/2015/>.
- 39 Carvalhais, N., and Coauthors, 2014: Global covariation of carbon turnover times with climate in terrestrial
40 ecosystems. *Nature*, **514**, 213–217, doi:10.1038/nature13731.
- 41 Caussen, M., V. Brovkin, and A. Ganopolski, 2001: Biophysical versus biogeochemical feedbacks of large-
42 scale land cover change. *Geophys. Res. Lett.*, **28**, 1011–1014, doi:10.1029/2000GL012471.
43 <http://onlinelibrary.wiley.com/doi/10.1029/2000GL012471/full>.
- 44 Cavan, G., and Coauthors, 2014: Urban morphological determinants of temperature regulating ecosystem
45 services in two African cities. *Ecol. Indic.*, **42**, 43–57,
46 doi:https://doi.org/10.1016/j.ecolind.2014.01.025.
47 <http://www.sciencedirect.com/science/article/pii/S1470160X14000338>.
- 48 Ceccherini, G., S. Russo, I. Amezttoy, C. Patricia Romero, and C. Carmona-Moreno, 2016: Magnitude and
49 frequency of heat and cold waves in recent decades: The case of South America. *Nat. Hazards Earth
50 Syst. Sci.*, **16**, 821–831, doi:10.5194/nhess-16-821-2016. [https://www.nat-hazards-earth-syst-
51 sci.net/16/821/2016/](https://www.nat-hazards-earth-syst-sci.net/16/821/2016/) (Accessed October 28, 2018).
- 52 —, —, —, A. Francesco Marchese, and C. Carmona-Moreno, 2017: Heat waves in Africa 1981–2015,
53 observations and reanalysis. *Nat. Hazards Earth Syst. Sci.*, **17**, 115–125, doi:10.5194/nhess-17-115-
54 2017.
- 55 Ceschia, E., B. Mary, M. Ferlicoq, G. Pique, D. Carrer, J. F. Dejoux, and G. Dedieu, 2017: Potentiel
56 d'atténuation des changements climatiques par les couverts intermédiaires. *Innov. Agron.*, **62**, 17–32.

- 1 Chadwick, D. R., and Coauthors, 2018: The contribution of cattle urine and dung to nitrous oxide emissions:
2 Quantification of country specific emission factors and implications for national inventories. *Sci. Total*
3 *Environ.*, **635**, 607–617, doi:10.1016/j.scitotenv.2018.04.152.
4 <https://doi.org/10.1016/j.scitotenv.2018.04.152>.
- 5 Chagas, V. B. P., and P. L. B. Chaffe, 2018: The Role of Land Cover in the Propagation of Rainfall Into
6 Streamflow Trends. *Water Resour. Res.*, **54**, 5986–6004, doi:10.1029/2018WR022947.
7 <http://doi.wiley.com/10.1029/2018WR022947> (Accessed March 25, 2019).
- 8 Challinor, A. J., J. Watson, D. B. Lobell, S. M. Howden, D. R. Smith, and N. Chhetri, 2014: A meta-analysis
9 of crop yield under climate change and adaptation. *Nat. Clim. Chang.*, **4**, 287–291,
10 doi:10.1038/nclimate2153.
- 11 Chan, D., and Q. Wu, 2015: Significant anthropogenic-induced changes of climate classes since 1950. *Sci.*
12 *Rep.*, **5**, 13487, doi:10.1038/srep13487. <http://www.nature.com/articles/srep13487> (Accessed April 15,
13 2019).
- 14 Chang, J., P. Ciais, N. Viovy, J. F. Soussana, K. Klumpp, and B. Sultan, 2017: Future productivity and
15 phenology changes in European grasslands for different warming levels: Implications for grassland
16 management and carbon balance. *Carbon Balance Manag.*, **12**, 11, doi:10.1186/s13021-017-0079-8.
17 <https://cbmjournal.biomedcentral.com/articles/10.1186/s13021-017-0079-8> (Accessed March 13,
18 2019).
- 19 Charney, J. G., 1975: Dynamics of deserts and drought in the Sahel. *Q. J. R. Meteorol. Soc.*, **101**, 193–202,
20 doi:10.1002/qj.49710142802.
- 21 Chase, T. N., R. A. Pielke, T. G. F. Kittel, R. R. Nemani, and S. W. Running, 2000: Simulated impacts of
22 historical land cover changes on global climate in northern winter. *Clim. Dyn.*,
23 doi:10.1007/s003820050007.
- 24 Chase, T. N., R. a. Pielke, T. G. F. Kittel, M. Zhao, a. J. Pitman, S. W. Running, and R. R. Nemani, 2001:
25 Relative climatic effects of landcover change and elevated carbon dioxide combined with aerosols: A
26 comparison of model results and observations. *J. Geophys. Res.*, **106**, 31685,
27 doi:10.1029/2000JD000129.
- 28 Chaudhary, N., P. A. Miller, and B. Smith, 2017: Modelling Holocene peatland dynamics with an individual-
29 based dynamic vegetation model. *Biogeosciences*, doi:10.5194/bg-14-2571-2017.
- 30 Chawla, I., K. K. Osuri, P. P. Mujumdar, and D. Niyogi, 2018: Assessment of the Weather Research and
31 Forecasting (WRF) model for simulation of extreme rainfall events in the upper Ganga Basin. *Hydrol.*
32 *Earth Syst. Sci.*, **22**, 1095–1117, doi:10.5194/hess-22-1095-2018. [https://www.hydrol-earth-syst-](https://www.hydrol-earth-syst-sci.net/22/1095/2018/)
33 [sci.net/22/1095/2018/](https://www.hydrol-earth-syst-sci.net/22/1095/2018/) (Accessed October 30, 2018).
- 34 Cheesman, S., C. Thierfelder, N. S. Eash, G. T. Kassie, and E. Frossard, 2016: Soil carbon stocks in
35 conservation agriculture systems of Southern Africa. *Soil Tillage Res.*, **156**, 99–109,
36 doi:10.1016/j.still.2015.09.018.
- 37 Chen, B., L. Dong, X. Liu, G. Y. Shi, L. Chen, T. Nakajima, and A. Habib, 2016a: Exploring the possible
38 effect of anthropogenic heat release due to global energy consumption upon global climate: a climate
39 model study. *Int. J. Climatol.*, **36**, 4790–4796, doi:10.1002/joc.4669.
- 40 Chen, C., and Coauthors, 2019: China and India lead in greening of the world through land-use management.
41 *Nat. Sustain.*, **2**, 122–129, doi:10.1038/s41893-019-0220-7. [http://www.nature.com/articles/s41893-](http://www.nature.com/articles/s41893-019-0220-7)
42 [019-0220-7](http://www.nature.com/articles/s41893-019-0220-7) (Accessed March 12, 2019).
- 43 Chen, D., and H. W. Chen, 2013: Using the Köppen classification to quantify climate variation and change:
44 An example for 1901–2010. *Environ. Dev.*, **6**, 69–79, doi:10.1016/j.envdev.2013.03.007.
45 <https://www.sciencedirect.com/science/article/pii/S2211464513000328> (Accessed April 15, 2019).
- 46 Chen, G. S., M. Notaro, Z. Liu, and Y. Liu, 2012: Simulated local and remote biophysical effects of
47 afforestation over the Southeast United States in boreal summer. *J. Clim.*, **25**, 4511–4522,
48 doi:10.1175/JCLI-D-11-00317.1.
- 49 Chen, J., 2007: Rapid urbanization in China: A real challenge to soil protection and food security. *CATENA*,
50 **69**, 1–15, doi:<https://doi.org/10.1016/j.catena.2006.04.019>.
51 <http://www.sciencedirect.com/science/article/pii/S0341816206000920>.
- 52 Chen, L., P. A. Dirmeyer, Z. Guo, and N. M. Schultz, 2018: Pairing FLUXNET sites to validate model
53 representations of land-use/land-cover change. *Hydrol. Earth Syst. Sci.*, **22**, 111–125, doi:10.5194/hess-
54 22-111-2018.
- 55 Chen, W., Y. Zhang, X. Mai, and Y. Shen, 2016b: Multiple mechanisms contributed to the reduced stability
56 of Inner Mongolia grassland ecosystem following nitrogen enrichment. *Plant Soil*, **409**, 283–296,

- 1 doi:10.1007/s11104-016-2967-1. <http://dx.doi.org/10.1007/s11104-016-2967-1>.
- 2 Chen, X., and S. J. Jeong, 2018: Irrigation enhances local warming with greater nocturnal warming effects
3 than daytime cooling effects. *Environ. Res. Lett.*, **13**, 24005, doi:10.1088/1748-9326/aa9dea.
4 <http://stacks.iop.org/1748-9326/13/i=2/a=024005?key=crossref.4a93ef61f758279985613f5e49fa2cd5>.
- 5 Chen, Y., D. C. Morton, N. Andela, G. R. Van Der Werf, L. Giglio, and J. T. Randerson, 2017: A pan-
6 tropical cascade of fire driven by El Niño/Southern Oscillation. *Nat. Clim. Chang.*, **7**, 906–911,
7 doi:10.1038/s41558-017-0014-8.
- 8 Cheng, L., and Coauthors, 2017: Warming enhances old organic carbon decomposition through altering
9 functional microbial communities. *Isme J.*, **11**, 1825. <https://doi.org/10.1038/ismej.2017.48>.
- 10 Cherubini, F., M. Huijbregts, G. Kindermann, R. Van Zelm, M. Van Der Velde, K. Stadler, and A. H.
11 Strømman, 2016: Global spatially explicit CO₂ emission metrics for forest bioenergy. *Sci. Rep.*, **6**,
12 20186, doi:10.1038/srep20186 [http://www.nature.com/articles/srep20186](http://www.nature.com/articles/srep20186#supplementary-information)
13 <http://dx.doi.org/10.1038/srep20186>.
- 14 Chiang, F., O. Mazdiyasn, and A. AghaKouchak, 2018: Amplified warming of droughts in southern United
15 States in observations and model simulations. *Sci. Adv.*, **4**, eaat2380, doi:10.1126/sciadv.aat2380.
- 16 Chitra-Tarak, R., L. Ruiz, H. S. Dattaraja, M. S. Mohan Kumar, J. Riotte, H. S. Suresh, S. M. McMahon, and
17 R. Sukumar, 2017: The roots of the drought: Hydrology and water uptake strategies mediate forest-
18 wide demographic response to precipitation. *J. Ecol.*, **106**, 1495–1507, doi:10.1111/1365-2745.12925.
- 19 Choat, B., and Coauthors, 2012: Global convergence in the vulnerability of forests to drought. *Nature*, **491**,
20 752–755, doi:10.1038/nature11688.
- 21 Cholakian, A., A. Colette, G. Ciarelli, I. Coll, and M. Beekmann, 2018: Future climatic drivers and their
22 effect on PM_{2.5} components in Europe and the
23 Mediterranean Sea. *Atmos. Chem. Phys. Discuss.*, 1–35, doi:10.5194/acp-2018-868.
24 <https://www.atmos-chem-phys-discuss.net/acp-2018-868/>.
- 25 Christen, A., R. Jassal, T. Black, N. Grant, I. Hawthorne, M. Johnson, S. Lee, and M. Merkens, 2016:
26 Summertime greenhouse gas fluxes from an urban bog undergoing restoration through rewetting. *Mires
27 Peat*, **17**, 1–24, doi:10.19189/MaP.2015.OMB.207. <http://www.mires-and-peat.net/>.
- 28 Christensen, J. H., and O. B. Christensen, 2003: Severe summertime flooding in Europe. *Nature*, **421**, 805–
29 806, doi:10.1038/421805a.
- 30 Christensen, J. H., and Coauthors, 2013: Climate phenomena and their relevance for future regional climate
31 change. *Climate Change 2013 the Physical Science Basis: Working Group I Contribution to the Fifth
32 Assessment Report of the Intergovernmental Panel on Climate Change*, T.F. Stocker et al., Eds., Vol.
33 9781107057 of, Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA,
34 1217–1308.
- 35 Christidis, N., P. A. Stott, G. C. Hegerl, and R. A. Betts, 2013: The role of land use change in the recent
36 warming of daily extreme temperatures. *Geophys. Res. Lett.*, **40**, 589–594, doi:10.1002/grl.50159.
- 37 Christopher, S. F., R. Lal, and U. Mishra, 2009: Regional study of no-till effects on carbon sequestration in
38 the Midwestern United States. *Soil Sci. Soc. Am. J.*, **73**, 207–216, doi:10.2136/sssaj2007.0336.
- 39 Christy, J. R., W. B. Norris, K. Redmond, and K. P. Gallo, 2006: Methodology and results of calculating
40 central California surface temperature trends: Evidence of human-induced climate change? *J. Clim.*, **19**,
41 548–563, doi:10.1175/JCLI3627.1.
- 42 Chrysanthou, A., G. Van Der Schrier, E. J. M. Van Den Besselaar, A. M. G. Klein Tank, and T. Brandsma,
43 2014: The effects of urbanization on the rise of the European temperature since 1960. *Geophys. Res.
44 Lett.*, **41**, 7716–7722, doi:10.1002/2014GL061154.
- 45 Chum, H., and Coauthors, 2011: Bioenergy. *IPCC Special Report on Renewable Energy Sources and
46 Climate Change Mitigation*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge, United
47 Kingdom and New York, NY, USA, United Kingdom and New York, NY, USA, 209–332.
- 48 Ciais, P., and Coauthors, 2005: Europe-wide reduction in primary productivity caused by the heat and
49 drought in 2003. *Nature*, **437**, 529–533, doi:10.1038/nature03972.
50 <http://www.nature.com/articles/nature03972> (Accessed October 30, 2018).
- 51 Ciais, P., and Coauthors, 2013a: Carbon and other biogeochemical cycles. *Climate Change 2013 the
52 Physical Science Basis: Working Group I Contribution to the Fifth Assessment Report of the
53 Intergovernmental Panel on Climate Change*, V.B. and P.M.M. (eds.). [Stocker, T.F., D. Qin, G.-K.
54 Plattner, M. Tignor, S.K. Allen, J. Boschung, A. Nauels, Y. Xia, Ed., Vol. 9781107057 of, IPCC, New
55 York, 465–570.
- 56 —, and Coauthors, 2013b: Chapter 6: Carbon and Other Biogeochemical Cycles. *Climate Report, Change*

- 1 2013: *The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment of the*
2 *Intergovernmental Panel on Climate Change*, T.F. Stocker et al., Eds., Cambridge University Press,
3 Cambridge, United Kingdom.
- 4 Cintas, O., G. Berndes, A. L. Cowie, G. Egnell, H. Holmström, G. Marland, and G. I. Ågren, 2017: Carbon
5 balances of bioenergy systems using biomass from forests managed with long rotations: bridging the
6 gap between stand and landscape assessments. *GCB Bioenergy*, **9**, 1238–1251,
7 doi:10.1111/gcbb.12425.
- 8 Cirino, G. G., R. A. F. Souza, D. K. Adams, and P. Artaxo, 2014: The effect of atmospheric aerosol particles
9 and clouds on net ecosystem exchange in the Amazon. *Atmos. Chem. Phys.*, **14**, 6523–6543,
10 doi:10.5194/acp-14-6523-2014.
- 11 Claeys, M., W. Wang, A. C. Ion, I. Kourtchev, A. Gelencsér, and W. Maenhaut, 2004: Formation of
12 secondary organic aerosols from isoprene and its gas-phase oxidation products through reaction with
13 hydrogen peroxide. *Atmos. Environ.*, **38**, 4093–4098, doi:10.1016/j.atmosenv.2004.06.001.
14 [http://linkinghub.elsevier.com/retrieve/pii/S1352231004004996%5Cn%3CGo to](http://linkinghub.elsevier.com/retrieve/pii/S1352231004004996%5Cn%3CGo%20to%20ISI%3E%2F000222706000001%5Cnhttp://ac.els-cdn.com/S1352231004004996/1-s2.0-S1352231004004996-main.pdf?_tid=92b72d1c-7be1-11e5-be24-00000aacb35f&acdnat=1445864626_a476e4a5b538fa72632)
15 [ISI%3E://000222706000001%5Cnhttp://ac.els-cdn.com/S1352231004004996/1-s2.0-](http://ac.els-cdn.com/S1352231004004996/1-s2.0-S1352231004004996-main.pdf?_tid=92b72d1c-7be1-11e5-be24-00000aacb35f&acdnat=1445864626_a476e4a5b538fa72632)
16 [S1352231004004996-main.pdf?_tid=92b72d1c-7be1-11e5-be24-](http://ac.els-cdn.com/S1352231004004996/1-s2.0-S1352231004004996-main.pdf?_tid=92b72d1c-7be1-11e5-be24-00000aacb35f&acdnat=1445864626_a476e4a5b538fa72632)
17 [00000aacb35f&acdnat=1445864626_a476e4a5b538fa72632.](http://ac.els-cdn.com/S1352231004004996/1-s2.0-S1352231004004996-main.pdf?_tid=92b72d1c-7be1-11e5-be24-00000aacb35f&acdnat=1445864626_a476e4a5b538fa72632)
- 18 Claussen, M., and Coauthors, 2001a: Biophysical versus biogeochemical feedbacks of large-scale land cover
19 change. *Geophys. Res. Lett.*, **28**, 1011–1014, doi:10.1029/2000GL012471.
- 20 —, V. Brovkin, and A. Ganopolski, 2001b: Biophysical versus biogeochemical feedbacks of large-scale
21 land cover change. *Geophys. Res. Lett.*, **28**, 1011–1014, doi:10.1029/2000GL012471.
- 22 Cleveland, C., K. Todd-brown, W. R. Wieder, C. C. Cleveland, W. K. Smith, and K. Todd-brown, 2015:
23 Future productivity and carbon storage limited by terrestrial nutrient availability Future productivity
24 and carbon storage limited by terrestrial nutrient availability. *Nat. Geosci.*, **8**, 441,
25 doi:10.1038/NNGEO2413. <http://dx.doi.org/10.1038/ngeo2413>.
- 26 Cohen, J. B., and C. Wang, 2014: Estimating global black carbon emissions using a top-down Kalman Filter
27 approach. *J. Geophys. Res. Atmos.*, **119**, 307–323, doi:10.1002/2013JD019912.
28 <http://doi.wiley.com/10.1002/2013JD019912>.
- 29 Colwell, R. K., G. Brehm, C. L. Cardelús, A. C. Gilman, and J. T. Longino, 2008a: Global warming,
30 elevational range shifts, and lowland biotic attrition in the wet tropics. *Science (80-.)*, **322**, 258–261,
31 doi:10.1126/science.1162547. <http://science.sciencemag.org/content/344/6183/1247579.abstract>.
- 32 —, —, —, —, and —, 2008b: Global warming, elevational range shifts, and lowland biotic
33 attrition in the wet tropics. *Science (80-.)*, **322**, 258–261, doi:10.1126/science.1162547.
- 34 Coma, J., G. Pérez, A. de Gracia, S. Burés, M. Urrestarazu, and L. F. Cabeza, 2017: Vertical greenery
35 systems for energy savings in buildings: A comparative study between green walls and green facades.
36 *Build. Environ.*, **111**, 228–237, doi:<https://doi.org/10.1016/j.buildenv.2016.11.014>.
- 37 Conant, R. T., and Coauthors, 2011: Temperature and soil organic matter decomposition rates - synthesis of
38 current knowledge and a way forward. *Glob. Chang. Biol.*, **17**, 3392–3404, doi:10.1111/j.1365-
39 2486.2011.02496.x. <https://doi.org/10.1111/j.1365-2486.2011.02496.x>.
- 40 —, C. E. P. Cerri, B. B. Osborne, and K. Paustian, 2017: Grassland management impacts on soil carbon
41 stocks: A new synthesis: A. *Ecol. Appl.*, doi:10.1002/eap.1473.
- 42 de Coninck, H., and Coauthors, 2018: Strengthening and implementing the global response. *Global Warming*
43 *of 1.5C: an IPCC special report on the impacts of global warming of 1.5C above pre-industrial levels*
44 *and related global greenhouse gas emission pathways, in the context of strengthening the global*
45 *response to the threat of climate change*, V. Masson-Delmotte et al., Eds., Geneva
46 <http://www.ipcc.ch/report/sr15/>.
- 47 Connor, E. W., and C. V. Hawkes, 2018: Effects of extreme changes in precipitation on the physiology of C4
48 grasses. *Oecologia*, **188**, 355–365, doi:10.1007/s00442-018-4212-5.
49 <http://link.springer.com/10.1007/s00442-018-4212-5> (Accessed November 2, 2018).
- 50 Constantin, J., B. Mary, F. Laurent, G. Aubrion, A. Fontaine, P. Kerveillant, and N. Beaudoin, 2010: Effects
51 of catch crops, no till and reduced nitrogen fertilization on nitrogen leaching and balance in three long-
52 term experiments. *Agric. Ecosyst. Environ.*, **135**, 268–278,
53 doi:<https://doi.org/10.1016/j.agee.2009.10.005>.
54 <http://www.sciencedirect.com/science/article/pii/S0167880909003089>.
- 55 Constantin, J., N. Beaudoin, F. Laurent, J. P. Cohan, F. Duyme, and B. Mary, 2011: Cumulative effects of
56 catch crops on nitrogen uptake, leaching and net mineralization. *Plant Soil*, **341**, 137–154,

- 1 doi:10.1007/s11104-010-0630-9.
- 2 Contini, D., R. Vecchi, and M. Viana, 2018: Carbonaceous Aerosols in the Atmosphere. *Atmosphere*
- 3 (*Basel*), **9**, 181, doi:10.3390/atmos9050181. <http://www.mdpi.com/2073-4433/9/5/181>.
- 4 Conway, G., 2012: How to create resilient agriculture. *Appropriate Technology*, O. Edenhofer et al., Eds.,
- 5 Vol. 39 of, Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 12–
- 6 14.
- 7 Cook, B. I., T. R. Ault, and J. E. Smerdon, 2015: Unprecedented 21st century drought risk in the American
- 8 Southwest and Central Plains. *Sci. Adv.*, **1**, e1400082–e1400082, doi:10.1126/sciadv.1400082.
- 9 ———, J. S. Mankin, and K. J. Anchukaitis, 2018: Climate Change and Drought: From Past to Future. *Curr.*
- 10 *Clim. Chang. Reports*, **4**, 164–179, doi:10.1007/s40641-018-0093-2.
- 11 da Costa, A. C. L., and Coauthors, 2010: Effect of 7 yr of experimental drought on vegetation dynamics and
- 12 biomass storage of an eastern Amazonian rainforest. *New Phytol.*, **187**, 579–591, doi:10.1111/j.1469-
- 13 8137.2010.03309.x. <http://doi.wiley.com/10.1111/j.1469-8137.2010.03309.x> (Accessed May 21, 2018).
- 14 Costa, H., G. Floater, H. Hooyberghs, W. P. No, and W. P. No, 2016: Climate change , heat stress and labour
- 15 productivity : A cost methodology for city economies July 2016 Centre for Climate Change Economics
- 16 and Policy the Environment.
- 17 Coumou, D., and S. Rahmstorf, 2012: A decade of weather extremes. *Nat. Clim. Chang.*, **2**, 491–496,
- 18 doi:10.1038/nclimate1452. <http://www.nature.com/doi/10.1038/nclimate1452> (Accessed May 19,
- 19 2018).
- 20 Couwenberg, J., R. Dommain, and H. Joosten, 2010: Greenhouse gas fluxes from tropical peatlands in south-
- 21 east Asia. *Glob. Chang. Biol.*, **16**, 1715–1732, doi:10.1111/j.1365-2486.2009.02016.x.
- 22 <https://doi.org/10.1111/j.1365-2486.2009.02016.x>.
- 23 Cowan, T., A. Purich, S. Perkins, A. Pezza, G. Boschat, and K. Sadler, 2014: More frequent, longer, and
- 24 hotter heat waves for Australia in the Twenty-First Century. *J. Clim.*, **27**, 5851–5871,
- 25 doi:10.1175/JCLI-D-14-00092.1. <http://journals.ametsoc.org/doi/abs/10.1175/JCLI-D-14-00092.1>
- 26 (Accessed October 28, 2018).
- 27 Cowie AL, Kirschbaum MUF, and Ward M., 2007: Options for including all lands in a future greenhouse
- 28 gas accounting framework. *Env. Sci Policy* , **10**, 306–321.
- 29 Cowling, S. A., C. D. Jones, and P. M. Cox, 2009: Greening the terrestrial biosphere: Simulated feedbacks
- 30 on atmospheric heat and energy circulation. *Clim. Dyn.*, **32**, 287–299, doi:10.1007/s00382-008-0481-8.
- 31 Cox, P. M., D. Pearson, B. B. Booth, P. Friedlingstein, C. Huntingford, C. D. Jones, and C. M. Luke, 2013:
- 32 Sensitivity of tropical carbon to climate change constrained by carbon dioxide variability. *Nature*, **494**,
- 33 341–344, doi:10.1038/nature11882.
- 34 Craine, J. M., and Coauthors, 2018: Isotopic evidence for oligotrophication of terrestrial ecosystems. *Nat.*
- 35 *Ecol. Evol.*, **2**, 1735–1744, doi:10.1038/s41559-018-0694-0.
- 36 Crétat, J., B. Pohl, Y. Richard, and P. Drobinski, 2012: Uncertainties in simulating regional climate of
- 37 Southern Africa: Sensitivity to physical parameterizations using WRF. *Clim. Dyn.*, **38**, 613–634,
- 38 doi:10.1007/s00382-011-1055-8. <http://link.springer.com/10.1007/s00382-011-1055-8> (Accessed
- 39 October 30, 2018).
- 40 Creutzig, F., and Coauthors, 2015a: Bioenergy and climate change mitigation: An assessment. *GCB*
- 41 *Bioenergy*, **7**, 916–944, doi:10.1111/gcbb.12205. <https://doi.org/10.1111/gcbb.12205>.
- 42 ———, G. Baiocchi, R. Bierkandt, P.-P. Pichler, and K. C. Seto, 2015b: Global typology of urban energy use
- 43 and potentials for an urbanization mitigation wedge. *Proc. Natl. Acad. Sci.*, **112**, 6283–6288,
- 44 doi:10.1073/pnas.1315545112.
- 45 Crimp, S. J., D. Gobbett, P. Kokic, U. Nidumolu, M. Howden, and N. Nicholls, 2016: Recent seasonal and
- 46 long-term changes in southern Australian frost occurrence. *Clim. Change*, **139**, 115–128,
- 47 doi:10.1007/s10584-016-1763-5.
- 48 Crook, J. A., L. S. Jackson, S. M. Osprey, and P. M. Forster, 2015: A comparison of temperature and
- 49 precipitation responses to different earth radiation management geoengineering schemes. *J. Geophys.*
- 50 *Res.*, **120**, 9352–9373, doi:10.1002/2015JD023269.
- 51 Crooks, S., D. Herr, J. Tamelander, D. Laffoley, and J. Vandever, 2011: *Mitigating climate change through*
- 52 *restoration and management of coastal wetlands and near-shore marine ecosystems : challenges and*
- 53 *opportunities*. 69 pp.
- 54 Crowther, T. W., S. M. Thomas, D. S. Maynard, P. Baldrian, K. Covey, S. D. Frey, L. T. A. van Diepen, and
- 55 M. A. Bradford, 2015: Biotic interactions mediate soil microbial feedbacks to climate change. *Proc.*
- 56 *Natl. Acad. Sci.*, **112**, 7033–7038, doi:10.1073/pnas.1502956112.

- 1 Crowther, T. W., and Coauthors, 2016: Quantifying global soil carbon losses in response to warming.
2 *Nature*, **540**, 104–108, doi:10.1038/nature20150. <http://dx.doi.org/10.1038/nature20150>.
- 3 Curry, C. L., 2009: The consumption of atmospheric methane by soil in a simulated future climate.
4 *Biogeosciences*, **6**, 2355–2367, doi:10.5194/bgd-6-6077-2009.
- 5 D’Amato, G., L. Cecchi, M. D’Amato, and G. Liccardi, 2010: Urban air pollution and climate change as
6 environmental risk factors of respiratory allergy: An update. *J. Investig. Allergol. Clin. Immunol.*, **20**,
7 95–102.
- 8 Dahiya, B., 2012: Cities in Asia, 2012: Demographics, economics, poverty, environment and governance.
9 *Cities*, **29**, S44–S61, doi:10.1016/j.cities.2012.06.013. <http://dx.doi.org/10.1016/j.cities.2012.06.013>.
- 10 Dai, A., 2011: Drought under global warming: A review. *Wiley Interdiscip. Rev. Clim. Chang.*, **2**, 45–65,
11 doi:10.1002/wcc.81.
- 12 ———, 2013: Increasing drought under global warming in observations and models. *Nat. Clim. Chang.*, **3**, 52–
13 58, doi:10.1038/nclimate1633. <http://www.nature.com/articles/nclimate1633> (Accessed December 20,
14 2017).
- 15 ———, and T. Zhao, 2017: Uncertainties in historical changes and future projections of drought. Part I:
16 estimates of historical drought changes. *Clim. Change*, **144**, 519–533, doi:10.1007/s10584-016-1705-2.
- 17 ———, ———, and J. Chen, 2018: Climate Change and Drought: a Precipitation and Evaporation Perspective.
18 *Curr. Clim. Chang. Reports*, **4**, 301–312, doi:10.1007/s40641-018-0101-6.
- 19 Daioglou, V., E. Stehfest, B. Wicke, A. Faaij, and D. P. van Vuuren, 2016: Projections of the availability and
20 cost of residues from agriculture and forestry. *GCB Bioenergy*, doi:10.1111/gcbb.12285.
- 21 ———, J. C. Doelman, E. Stehfest, C. Müller, B. Wicke, A. Faaij, and D. P. Van Vuuren, 2017: Greenhouse
22 gas emission curves for advanced biofuel supply chains. *Nat. Clim. Chang.*, doi:10.1038/s41558-017-
23 0006-8.
- 24 ———, ———, B. Wicke, A. Faaij, and D. P. van Vuuren, 2019: Integrated assessment of biomass supply and
25 demand in climate change mitigation scenarios. *Glob. Environ. Chang.*,
26 doi:10.1016/j.gloenvcha.2018.11.012.
- 27 Dale, V. H., E. Parish, K. L. Kline, and E. Tobin, 2017: How is wood-based pellet production affecting
28 forest conditions in the southeastern United States? *For. Ecol. Manage.*,
29 doi:10.1016/j.foreco.2017.03.022.
- 30 Dangal, S. R. S., and Coauthors, 2019: Global Nitrous Oxide Emissions From Pasturelands and Rangelands:
31 Magnitude, Spatiotemporal Patterns, and Attribution. *Global Biogeochem. Cycles*, **33**, 200–222,
32 doi:10.1029/2018GB006091. <http://doi.wiley.com/10.1029/2018GB006091>.
- 33 Daniels, E. E., G. Lenderink, R. W. A. Hutjes, and A. A. M. Holtslag, 2016: Observed urban effects on
34 precipitation along the Dutch West coast. *Int. J. Climatol.*, **36**, 2111–2119, doi:10.1002/joc.4458.
- 35 Dargie, G. C., and Coauthors, 2019: Congo Basin peatlands: threats and conservation priorities. *Mitig.*
36 *Adapt. Strateg. Glob. Chang.*, **24**, 669–686, doi:10.1007/s11027-017-9774-8.
- 37 Dass, P., C. Müller, V. Brovkin, and W. Cramer, 2013: Can bioenergy cropping compensate high carbon
38 emissions from large-scale deforestation of high latitudes? *Earth Syst. Dyn.*, **4**, 409–424,
39 doi:10.5194/esd-4-409-2013.
- 40 Davidson, E. A., 2009: The contribution of manure and fertilizer nitrogen to atmospheric nitrous oxide since
41 1860. *Nat. Geosci.*, **2**, 659–662, doi:10.1038/ngeo608. <http://dx.doi.org/10.1038/ngeo608>.
- 42 ———, I. A. Janssens, E. A. Davidson, and I. A. Janssens, 2006a: Temperature sensitivity of soil carbon
43 decomposition and feedbacks to climate change. *Nature*, **440**, 165–173, doi:10.1038/nature04514.
- 44 Davidson, E. A., I. A. Janssens, and Y. Q. Luo, 2006b: On the variability of respiration in terrestrial
45 ecosystems: moving beyond Q(10). *Glob. Chang. Biol.*, **12**, 154–164, doi:10.1111/j.1365-
46 2486.2005.01065.x.
- 47 Davidson, E. A., and Coauthors, 2012: The Amazon basin in transition. *Nature*, **481**, 321–328,
48 doi:10.1038/nature10717.
- 49 Davies-Barnard, T., P. J. Valdes, J. S. Singarayer, F. M. Pacifico, and C. D. Jones, 2014: Full effects of land
50 use change in the representative concentration pathways. *Environ. Res. Lett.*, **9**, doi:10.1088/1748-
51 9326/9/11/114014.
- 52 Davies-Barnard, T., P. J. Valdes, J. S. Singarayer, A. J. Wiltshire, and C. D. Jones, 2015: Quantifying the
53 relative importance of land cover change from climate and land use in the representative concentration
54 pathways. *Global Biogeochem. Cycles*, **29**, 842–853, doi:10.1002/2014GB004949.
55 <https://doi.org/10.1002/2014GB004949>.
- 56 Davies, Z. G., J. L. Edmondson, A. Heinemeyer, J. R. Leake, and K. J. Gaston, 2011: Mapping an urban

- 1 ecosystem service: Quantifying above-ground carbon storage at a city-wide scale. *J. Appl. Ecol.*, **48**,
2 1125–1134, doi:10.1111/j.1365-2664.2011.02021.x.
- 3 Davin, E. L., and N. de Noblet-Ducoudre, 2010: Climatic impact of global-scale Deforestation: Radiative
4 versus nonradiative processes. *J. Clim.*, **23**, 97–112, doi:10.1175/2009JCLI3102.1.
- 5 ———, N. de Noblet-Ducoudré, and P. Friedlingstein, 2007: Impact of land cover change on surface climate:
6 Relevance of the radiative forcing concept. *Geophys. Res. Lett.*, **34**, n/a-n/a,
7 doi:10.1029/2007GL029678.
- 8 ———, S. I. Seneviratne, P. Ciais, A. Orlino, and T. Wang, 2014: Preferential cooling of hot extremes from
9 cropland albedo management. *Proc. Natl. Acad. Sci.*, **111**, 9757–9761, doi:10.1073/pnas.1317323111.
10 <http://www.pnas.org/lookup/doi/10.1073/pnas.1317323111>.
- 11 Demuzere, M., and Coauthors, 2014: Mitigating and adapting to climate change: Multi-functional and multi-
12 scale assessment of green urban infrastructure. *J. Environ. Manage.*, **146**, 107–115,
13 doi:<https://doi.org/10.1016/j.jenvman.2014.07.025>.
14 <http://www.sciencedirect.com/science/article/pii/S0301479714003740>.
- 15 Derksen, C., and R. Brown, 2012: Spring snow cover extent reductions in the 2008–2012 period exceeding
16 climate model projections. *Geophys. Res. Lett.*, **39**, doi:10.1029/2012GL053387.
17 <https://doi.org/10.1029/2012GL053387>.
- 18 Derner, J. D., and Coauthors, 2003: Above- and below-ground responses of C3-C4 species mixtures to
19 elevated CO₂ and soil water availability. *Glob. Chang. Biol.*, **9**, 452–460, doi:10.1046/j.1365-
20 2486.2003.00579.x.
- 21 Deryng, D., and Coauthors, 2016a: Regional disparities in the beneficial effects of rising CO₂ concentrations
22 on crop water productivity. *Nat. Clim. Chang.*, doi:10.1038/nclimate2995.
- 23 ———, and Coauthors, 2016b: Regional disparities in the beneficial effects of rising CO₂ concentrations on
24 crop water productivity. *Nat. Clim. Chang.*, **6**, 786–790, doi:10.1038/nclimate2995.
- 25 Deser, C., A. Phillips, V. Bourdette, and H. Teng, 2012: 1. Uncertainty in climate change projections: The
26 role of internal variability. *Clim. Dyn.*, **38**, 527–546, doi:10.1007/s00382-010-0977-x.
27 <http://link.springer.com/10.1007/s00382-010-0977-x> (Accessed November 1, 2018).
- 28 Devaraju, N., G. Bala, and A. Modak, 2015a: Effects of large-scale deforestation on precipitation in the
29 monsoon regions: Remote versus local effects. *Proc. Natl. Acad. Sci.*, **112**, 3257–3262,
30 doi:10.1073/pnas.1423439112.
- 31 ———, ———, and ———, 2015b: Effects of large-scale deforestation on precipitation in the monsoon regions:
32 Remote versus local effects. *Proc. Natl. Acad. Sci.*, **112**, 3257–3262, doi:10.1073/pnas.1423439112.
- 33 Devaraju, N., G. Bala, and R. Nemani, 2015c: Modelling the influence of land-use changes on biophysical
34 and biochemical interactions at regional and global scales. *Plant, Cell Environ.*,
35 doi:10.1111/pce.12488.
- 36 Devaraju, N., G. Bala, K. Caldeira, and R. Nemani, 2016: A model based investigation of the relative
37 importance of CO₂-fertilization, climate warming, nitrogen deposition and land use change on the
38 global terrestrial carbon uptake in the historical period. *Clim. Dyn.*, doi:10.1007/s00382-015-2830-8.
- 39 Devaraju, N., N. de Noblet-Ducoudré, G. Bala, and B. Quesada, 2018a: Quantifying the relative importance
40 of direct and indirect biophysical effects of deforestation on surface temperature and teleconnections. *J.*
41 *Clim.*, **in rev.**
- 42 Devaraju, N., N. de Noblet-Ducoudré, B. Quesada, and G. Bala, 2018b: Quantifying the relative importance
43 of direct and indirect biophysical effects of deforestation on surface temperature and teleconnections. *J.*
44 *Clim.*, **in rev.**, JCLI-D-17-0563.1, doi:10.1175/JCLI-D-17-0563.1.
- 45 Dickie, A., C. Streck, S. Roe, M. Zurek, F. Haupt, and A. Dolginow, 2014: *Strategies for Mitigating Climate*
46 *Change in Agriculture*. 87 pp. www.agriculturalmitigation.org.
- 47 Diffenbaugh, N. S., and C. B. Field, 2013: Changes in ecologically critical terrestrial climate conditions.
48 *Science (80-.)*, **341**, 486–492, doi:10.1126/science.1237123.
49 <http://www.ncbi.nlm.nih.gov/pubmed/23908225> (Accessed March 12, 2019).
- 50 Diffenbaugh, N. S., D. L. Swain, and D. Touma, 2015: Anthropogenic warming has increased drought risk in
51 California. *Proc. Natl. Acad. Sci.*, **112**, 3931–3936, doi:10.1073/pnas.1422385112.
52 <http://www.ncbi.nlm.nih.gov/pubmed/25733875> (Accessed September 22, 2018).
- 53 Van Dijk, A. I. J. M., H. E. Beck, R. S. Crosbie, R. A. M. De Jeu, Y. Y. Liu, G. M. Podger, B. Timbal, and
54 N. R. Viney, 2013: The Millennium Drought in southeast Australia (2001–2009): Natural and human
55 causes and implications for water resources, ecosystems, economy, and society. *Water Resour. Res.*,
56 **49**, 1040–1057, doi:10.1002/wrcr.20123. <http://doi.wiley.com/10.1002/wrcr.20123> (Accessed May 7,

- 1 2018).
- 2 Dirmeyer, P. A., Z. Wang, M. J. Mbulu, and H. E. Norton, 2014: Intensified land surface control on boundary
3 layer growth in a changing climate. *Geophys. Res. Lett.*, **45**, 1290–1294,
4 doi:10.1002/2013GL058826. Received.
- 5 Dlugokencky, E. J., L. P. Steele, P. M. Lang, and K. a. Masarie, 1994: The growth rate and distribution of
6 atmospheric methane. *J. Geophys. Res.*, **99**, 17021, doi:10.1029/94JD01245.
7 <https://doi.org/10.1029/94JD01245>.
- 8 Doan, Q. Van, H. Kusaka, and Q. B. Ho, 2016: Impact of future urbanization on temperature and thermal
9 comfort index in a developing tropical city: Ho Chi Minh City. *Urban Clim.*, **17**, 20–31,
10 doi:10.1016/j.uclim.2016.04.003.
- 11 Doerr, S. H., and C. Santín, 2016: Global trends in wildfire and its impacts: Perceptions versus realities in a
12 changing world. *Philos. Trans. R. Soc. B Biol. Sci.*, **371**, 20150345, doi:10.1098/rstb.2015.0345.
- 13 Don, A., and Coauthors, 2012: Land-use change to bioenergy production in Europe: implications for the
14 greenhouse gas balance and soil carbon. *GCB Bioenergy*, **4**, 372–391, doi:10.1111/j.1757-
15 1707.2011.01116.x. <http://dx.doi.org/10.1111/j.1757-1707.2011.01116.x>.
- 16 Donat, M. G., A. J. Pitman, and O. Angéilil, 2018: Understanding and Reducing Future Uncertainty in
17 Midlatitude Daily Heat Extremes Via Land Surface Feedback Constraints. *Geophys. Res. Lett.*, **45**,
18 10,627–10,636, doi:10.1029/2018GL079128.
- 19 Dong, J., and W. T. Crow, 2019: L-band remote-sensing increases sampled levels of global soil moisture-air
20 temperature coupling strength. *Remote Sens. Environ.*, **220**, 51–58, doi:10.1016/j.rse.2018.10.024.
- 21 Donohue, R. J., M. L. Roderick, T. R. McVicar, and G. D. Farquhar, 2013: Impact of CO₂ fertilization on
22 maximum foliage cover across the globe's warm, arid environments. *Geophys. Res. Lett.*, **40**, 3031–
23 3035, doi:10.1002/grl.50563.
- 24 Dooley, K., and S. Kartha, 2018: Land-based negative emissions: Risks for climate mitigation and impacts
25 on sustainable development. *Int. Environ. Agreements Polit. Law Econ.*, doi:10.1007/s10784-017-
26 9382-9.
- 27 Dosio, A., 2017: Projection of temperature and heat waves for Africa with an ensemble of CORDEX
28 Regional Climate Models. *Clim. Dyn.*, **49**, 493–519, doi:10.1007/s00382-016-3355-5.
- 29 ———, L. Mentaschi, E. M. Fischer, and K. Wyser, 2018: Extreme heat waves under 1.5 °C and 2 °C global
30 warming. *Environ. Res. Lett.*, **13**, 54006, doi:10.1088/1748-9326/aab827.
- 31 Dou, J., Y. Wang, R. Bornstein, and S. Miao, 2014: Observed Spatial Characteristics of Beijing Urban
32 Climate Impacts on Summer Thunderstorms. *J. Appl. Meteorol. Climatol.*, **54**, 94–105,
33 doi:10.1175/JAMC-D-13-0355.1. <https://doi.org/10.1175/JAMC-D-13-0355.1>.
- 34 Doughty, C. E., and Coauthors, 2015: Drought impact on forest carbon dynamics and fluxes in Amazonia.
35 *Nature*, **519**, 78–82, doi:10.1038/nature14213. <http://www.nature.com/articles/nature14213> (Accessed
36 May 7, 2018).
- 37 Drake, J. E., and Coauthors, 2011: Increases in the flux of carbon belowground stimulate nitrogen uptake
38 and sustain the long-term enhancement of forest productivity under elevated CO₂. *Ecol. Lett.*, **14**, 349–
39 357, doi:10.1111/j.1461-0248.2011.01593.x. <https://doi.org/10.1111/j.1461-0248.2011.01593.x>.
- 40 Drösler, M., L. V. Verchot, A. Freibauer, and G. Pan, 2013: Drained inland organic soils. *2013 Suppl. to*
41 *2006 IPCC Guidel. Natl. Greenh. Gas Invent. Wetl.*, 1–79. [http://www.ipcc-](http://www.ipcc-nggip.iges.or.jp/public/wetlands/)
42 [nggip.iges.or.jp/public/wetlands/](http://www.ipcc-nggip.iges.or.jp/public/wetlands/).
- 43 Du, C., X. Wang, M. Zhang, J. Jing, and Y. Gao, 2019a: Effects of elevated CO₂ on plant C-N-P
44 stoichiometry in terrestrial ecosystems: A meta-analysis. *Sci. Total Environ.*, **650**, 697–708,
45 doi:10.1016/j.scitotenv.2018.09.051.
- 46 Du, C. J., X. D. Wang, M. Y. Zhang, J. Jing, and Y. H. Gao, 2019b: Effects of elevated CO₂ on plant C-N-P
47 stoichiometry in terrestrial ecosystems: A meta-analysis. *Sci. Total Environ.*, **650**, 697–708,
48 doi:10.1016/j.scitotenv.2018.09.051. [https://ac.els-cdn.com/S0048969718334818/1-s2.0-](https://ac.els-cdn.com/S0048969718334818/1-s2.0-S0048969718334818-main.pdf?_tid=b168b60e-67e8-4f67-a732-c53f23151654&acdnat=1540800035_cfe2cd673b638dea84429c306dc6390a)
49 [S0048969718334818-main.pdf?_tid=b168b60e-67e8-4f67-a732-](https://ac.els-cdn.com/S0048969718334818-main.pdf?_tid=b168b60e-67e8-4f67-a732-c53f23151654&acdnat=1540800035_cfe2cd673b638dea84429c306dc6390a)
50 [c53f23151654&acdnat=1540800035_cfe2cd673b638dea84429c306dc6390a](https://ac.els-cdn.com/S0048969718334818-main.pdf?_tid=b168b60e-67e8-4f67-a732-c53f23151654&acdnat=1540800035_cfe2cd673b638dea84429c306dc6390a).
- 51 Du, X., and Z. Huang, 2017: Ecological and environmental effects of land use change in rapid urbanization:
52 The case of Hangzhou, China. *Ecol. Indic.*, **81**, 243–251, doi:10.1016/j.ecolind.2017.05.040.
- 53 Duarte, C. M., 2017: Reviews and syntheses: Hidden forests, the role of vegetated coastal habitats in the
54 ocean carbon budget. *Biogeosciences*, **14**, 301–310, doi:10.5194/bg-14-301-2017.
55 <https://www.biogeosciences.net/14/301/2017/>.
- 56 Duarte, C. M., I. J. Losada, I. E. Hendriks, I. Mazarrasa, and N. Marba, 2013: The role of coastal plant

- 1 communities for climate change mitigation and adaptation. *Nat. Clim. Chang.*, **3**, 961–968,
2 doi:10.1038/nclimate1970 ([http://www.nature.com/nclimate/journal/v3/n11/abs/nclimate1970.html#su](http://www.nature.com/nclimate/journal/v3/n11/abs/nclimate1970.html#supplementary-information)
3 [pplementary-information. http://dx.doi.org/10.1038/nclimate1970](http://dx.doi.org/10.1038/nclimate1970)).
- 4 Duden, A. S., and Coauthors, 2017: Modeling the impacts of wood pellet demand on forest dynamics in
5 southeastern United States. *Biofuels, Bioprod. Biorefining*, doi:10.1002/bbb.1803.
- 6 Dungait, J. A. J., D. W. Hopkins, A. S. Gregory, and A. P. Whitmore, 2012: Soil organic matter turnover is
7 governed by accessibility not recalcitrance. *Glob. Chang. Biol.*, **18**, 1781–1796, doi:10.1111/j.1365-
8 2486.2012.02665.x.
- 9 Dutaur, L., and L. V. Verchot, 2007: A global inventory of the soil CH₄ sink. *Global Biogeochem. Cycles*,
10 **21**, 1–9, doi:10.1029/2006GB002734.
- 11 Dutta, M. K., T. S. Bianchi, and S. K. Mukhopadhyay, 2017: Mangrove Methane Biogeochemistry in the
12 Indian Sundarbans: A Proposed Budget. *Front. Mar. Sci.*, **4**, 187, doi:10.3389/fmars.2017.00187.
13 <http://journal.frontiersin.org/article/10.3389/fmars.2017.00187/full>.
- 14 Duveiller, G., J. Hooker, and A. Cescatti, 2018a: The mark of vegetation change on Earth's surface energy
15 balance. *Nat. Commun.*, **9**, 679, doi:10.1038/s41467-017-02810-8.
- 16 —, —, and —, 2018b: The mark of vegetation change on Earth's surface energy balance. *Nat.*
17 *Commun.*, **9**, 679, doi:10.1038/s41467-017-02810-8. <https://doi.org/10.1038/s41467-017-02810-8>.
- 18 Edmondson, J. L., Z. G. Davies, K. J. Gaston, and J. R. Leake, 2014: Urban cultivation in allotments
19 maintains soil qualities adversely affected by conventional agriculture. *J. Appl. Ecol.*, **51**, 880–889,
20 doi:10.1111/1365-2664.12254.
- 21 Edwards, M. E., L. B. Brubaker, A. V. Lozhkin, and M. Patricia, 2018: Structurally Novel Biomes : A
22 Response to past Warming in Beringia Published by : Wiley on behalf of the Ecological Society of
23 America Stable URL : <http://www.jstor.org/stable/3450612> REFERENCES Linked references are
24 available on JSTOR for this article : . **86**, 1696–1703.
- 25 Eekhout, J. P. C., J. E. Hunink, W. Terink, and J. de Vente, 2018: Why increased extreme precipitation under
26 climate change negatively affects water security. *Hydrol. Earth Syst. Sci. Discuss.*, 1–16,
27 doi:10.5194/hess-2018-161. <https://www.hydrol-earth-syst-sci-discuss.net/hess-2018-161/> (Accessed
28 September 23, 2018).
- 29 Elagib, N. A., 2011: Evolution of urban heat island in Khartoum. *Int. J. Climatol.*, **31**, 1377–1388,
30 doi:10.1002/joc.2159.
- 31 Elbert, W., B. Weber, S. Burrows, J. Steinkamp, B. Büdel, M. O. Andreae, and U. Pöschl, 2012:
32 Contribution of cryptogamic covers to the global cycles of carbon and nitrogen. *Nat. Geosci.*, **5**, 459–
33 462, doi:10.1038/ngeo1486. <http://www.nature.com/articles/ngeo1486> (Accessed September 20, 2018).
- 34 Elliott, J., and Coauthors, 2014: Constraints and potentials of future irrigation water availability on
35 agricultural production under climate change. *Proc. Natl. Acad. Sci.*, **111**, 3239–3244,
36 doi:10.1073/pnas.1222474110. <http://www.pnas.org/lookup/doi/10.1073/pnas.1222474110>.
- 37 Ellison, D., and Coauthors, 2017a: Trees, forests and water: Cool insights for a hot world. *Glob. Environ.*
38 *Chang.*, **43**, 51–61, doi:10.1016/j.gloenvcha.2017.01.002.
39 <http://dx.doi.org/10.1016/j.gloenvcha.2017.01.002>.
- 40 —, and Coauthors, 2017b: Trees, forests and water: Cool insights for a hot world. *Glob. Environ. Chang.*,
41 **43**, 51–61, doi:10.1016/j.gloenvcha.2017.01.002.
- 42 Ellsworth, D. S., and Coauthors, 2017: Elevated CO₂ does not increase eucalypt forest productivity on a
43 low-phosphorus soil. *Nat. Clim. Chang.*, **7**, 279–282, doi:10.1038/nclimate3235.
44 <http://www.nature.com/doi/10.1038/nclimate3235>.
- 45 Elmar, K., and Coauthors, 2018: Pathways limiting warming to 1.5°C: a tale of turning around in no time?
46 *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, **376**, 20160457, doi:10.1098/rsta.2016.0457.
- 47 Elsen, P. R., and M. W. Tingley, 2015: Global mountain topography and the fate of montane species under
48 climate change. *Nat. Clim. Chang.*, **5**, 772–776, doi:10.1038/nclimate2656.
49 <http://www.nature.com/articles/nclimate2656> (Accessed October 31, 2018).
- 50 Elshout, P. M. F., R. van Zelm, J. Balkovic, M. Obersteiner, E. Schmid, R. Skalsky, M. van der Velde, and
51 M. A. J. Huijbregts, 2015: Greenhouse-gas payback times for crop-based biofuels. *Nat. Clim. Chang.*,
52 **5**, 604–610, doi:10.1038/nclimate2642
53 <http://www.nature.com/nclimate/journal/v5/n6/abs/nclimate2642.html#supplementary-information>.
54 <http://dx.doi.org/10.1038/nclimate2642>.
- 55 Engelbrecht, F., and Coauthors, 2015: Projections of rapidly rising surface temperatures over Africa under
56 low mitigation. *Environ. Res. Lett.*, **10**, 85004, doi:10.1088/1748-9326/10/8/085004.

- 1 <http://stacks.iop.org/1748-9326/10/i=8/a=085004?key=crossref.99c873fbf22c3a3a455a413c9e670e3c>
2 (Accessed March 11, 2019).
- 3 Environmental Protection Agency, 2015: *Stormwater Management in Response to Climate Change Impacts:
4 Lessons from the Chesapeake Bay and Great Lakes Regions (Final Report)*. , EPA/600/R-15/087F,
5 2015. Washington, DC, 80 pp. <https://cfpub.epa.gov/ncea/global/recordisplay.cfm?deid=310045>.
- 6 Epstein, H., and Coauthors, 2018: *Tundra greenness [in Arctic Report Card 2018]*,
7 <https://www.arctic.noaa.gov/Report-Card>. 46-52 pp.
- 8 Erb, K.-H., and Coauthors, 2016: *Biomass turnover time in terrestrial ecosystems halved by land use*.
9 ———, and Coauthors, 2017: Land management: data availability and process understanding for global
10 change studies. *Glob. Chang. Biol.*, **23**, 512–533, doi:10.1111/gcb.13443.
- 11 ———, and Coauthors, 2018a: Unexpectedly large impact of forest management and grazing on global
12 vegetation biomass. *Nature*, **553**, 73–76, doi:10.1038/nature25138. [https://mail-
13 attachment.googleusercontent.com/attachment/u/0/?ui=2&ik=03e7b0d8bd&attid=0.1&th=1607f69aea2
14 25672&view=att&disp=inline&sadnir=1&saddbat=ANGjdJ9CDan5p6BLIbdOTgzuhFb4s5ubaWK_I7
15 GRvgjix-1BFJdJqGIgYVkj5jtWcZMDxelV3jtvsv77_dCNs0Go4vu6yTtRXDgLhlRvj82](https://mail-attachment.googleusercontent.com/attachment/u/0/?ui=2&ik=03e7b0d8bd&attid=0.1&th=1607f69aea225672&view=att&disp=inline&sadnir=1&saddbat=ANGjdJ9CDan5p6BLIbdOTgzuhFb4s5ubaWK_I7GRvgjix-1BFJdJqGIgYVkj5jtWcZMDxelV3jtvsv77_dCNs0Go4vu6yTtRXDgLhlRvj82).
- 16 Erb, K. H., and Coauthors, 2018b: Unexpectedly large impact of forest management and grazing on global
17 vegetation biomass. *Nature*, **553**, 73–76, doi:10.1038/nature25138. [https://mail-
18 attachment.googleusercontent.com/attachment/u/0/?ui=2&ik=03e7b0d8bd&attid=0.1&th=1607f69aea2
19 25672&view=att&disp=inline&sadnir=1&saddbat=ANGjdJ9CDan5p6BLIbdOTgzuhFb4s5ubaWK_I7
20 GRvgjix-1BFJdJqGIgYVkj5jtWcZMDxelV3jtvsv77_dCNs0Go4vu6yTtRXDgLhlRvj82](https://mail-attachment.googleusercontent.com/attachment/u/0/?ui=2&ik=03e7b0d8bd&attid=0.1&th=1607f69aea225672&view=att&disp=inline&sadnir=1&saddbat=ANGjdJ9CDan5p6BLIbdOTgzuhFb4s5ubaWK_I7GRvgjix-1BFJdJqGIgYVkj5jtWcZMDxelV3jtvsv77_dCNs0Go4vu6yTtRXDgLhlRvj82).
- 21 Ercilla-Montserrat, M., P. Muñoz, J. I. Montero, X. Gabarrell, and J. Rieradevall, 2018: A study on air
22 quality and heavy metals content of urban food produced in a Mediterranean city (Barcelona). *J. Clean.
23 Prod.*, **195**, 385–395, doi:<https://doi.org/10.1016/j.jclepro.2018.05.183>.
24 <http://www.sciencedirect.com/science/article/pii/S0959652618315294>.
- 25 Erfanian, A., G. Wang, and L. Fomenko, 2017: Unprecedented drought over tropical South America in 2016:
26 Significantly under-predicted by tropical SST. *Sci. Rep.*, **7**, 5811, doi:10.1038/s41598-017-05373-2.
27 <http://www.nature.com/articles/s41598-017-05373-2> (Accessed November 1, 2018).
- 28 Ernakovich, J. G., E. J. Kachergis, R. T. Simpson, A. B. Berdanier, K. A. Hopping, M. D. Wallenstein, and
29 H. Steltzer, 2014: Predicted responses of arctic and alpine ecosystems to altered seasonality under
30 climate change. *Glob. Chang. Biol.*, **20**, 3256–3269, doi:10.1111/gcb.12568.
31 <http://doi.wiley.com/10.1111/gcb.12568> (Accessed October 31, 2018).
- 32 Erşahin, S., B. C. Bilgili, Ü. Dikmen, and I. Ercanli, 2016: Net primary productivity of anatolian forests in
33 relation to climate, 2000-2010. *For. Sci.*, **62**, 698–709, doi:10.5849/forsci.15-171.
34 <https://academic.oup.com/forestscience/article/62/6/698-709/4583975> (Accessed September 23, 2018).
- 35 Esch, T., and Coauthors, 2017: Breaking new ground in mapping human settlements from space – The
36 Global Urban Footprint. *ISPRS J. Photogramm. Remote Sens.*, **134**, 30–42,
37 doi:10.1016/j.isprsjprs.2017.10.012.
- 38 Espinoza, J. C., J. A. Marengo, J. Ronchail, J. M. Carpio, L. N. Flores, and J. L. Guyot, 2014: The extreme
39 2014 flood in south-western Amazon basin: The role of tropical-subtropical South Atlantic SST
40 gradient. *Environ. Res. Lett.*, **9**, 124007, doi:10.1088/1748-9326/9/12/124007.
41 <http://stacks.iop.org/1748-9326/9/i=12/a=124007?key=crossref.2af184a0ef227c5f43a2c65f42795911>
42 (Accessed March 24, 2019).
- 43 ———, J. Ronchail, J. A. Marengo, and H. Segura, 2018: Contrasting North–South changes in Amazon wet-
44 day and dry-day frequency and related atmospheric features (1981–2017). *Climate Dynamics*,
45 September 26 <http://link.springer.com/10.1007/s00382-018-4462-2> (Accessed March 20, 2019).
- 46 Evan, A. T., 2018: Surface Winds and Dust Biases in Climate Models. *Geophys. Res. Lett.*, **45**, 1079–1085,
47 doi:10.1002/2017GL076353.
- 48 Evan, A. T., C. Flamant, S. Fiedler, and O. Doherty, 2014: An analysis of aeolian dust in climate models.
49 *Geophys. Res. Lett.*, **41**, 5996–6001, doi:10.1002/2014GL060545.
50 <https://doi.org/10.1002/2014GL060545>.
- 51 ———, ———, M. Gaetani, and F. Guichard, 2016: The past, present and future of African dust. *Nature*, **531**,
52 493–495, doi:10.1038/nature17149. <http://dx.doi.org/10.1038/nature17149>.
- 53 Evans, S., P. Ginoux, S. Malyshev, and E. Shevliakova, 2016: Climate-vegetation interaction and
54 amplification of Australian dust variability. *Geophys. Res. Lett.*, **43**, 11,823-11,830,
55 doi:10.1002/2016GL071016. <http://doi.wiley.com/10.1002/2016GL071016> (Accessed January 23,
56 2018).

- 1 Fadrique, B., and Coauthors, 2018: Widespread but heterogeneous responses of Andean forests to climate
2 change. *Nature*, **564**, 207–212, doi:10.1038/s41586-018-0715-9.
3 <http://www.nature.com/articles/s41586-018-0715-9> (Accessed March 12, 2019).
- 4 Fahey, C., K. Winter, M. Slot, and K. Kitajima, 2016: Influence of arbuscular mycorrhizal colonization on
5 whole-plant respiration and thermal acclimation of tropical tree seedlings. *Ecol. Evol.*, **6**, 859–870,
6 doi:10.1002/ece3.1952.
- 7 Fanzo, J., C. Davis, R. McLaren, and J. Choufani, 2018: The effect of climate change across food systems:
8 Implications for nutrition outcomes. *Glob. Food Sec.*, **18**, 12–19, doi:10.1016/j.gfs.2018.06.001.
9 <https://www.sciencedirect.com/science/article/pii/S2211912418300063> (Accessed September 28,
10 2018).
- 11 FAO, 2015a: FAOSTAT. Land use emissions database. <http://faostat.fao.org/site/705/default.aspx>.
12 ———, 2015b: FAOSTAT- Emissions from land use. *Food Agric. Organ. United Nations.*,
13 FAO-FRA, 2015: *The Global Forest Resources Assessment 2015*. <http://www.fao.org/3/a-i4808e.pdf>.
- 14 Farrior, C. E., I. Rodriguez-Iturbe, R. Dybzinski, S. A. Levin, and S. W. Pacala, 2015: Decreased water
15 limitation under elevated CO₂ amplifies potential for forest carbon sinks. *Proc. Natl. Acad. Sci.*, **112**,
16 7213–7218, doi:10.1073/pnas.1506262112.
- 17 Favis-Mortlock, D., and D. Mullen, 2011: Soil erosion by water under future climate change. *Soil hydrology,*
18 *land use and agriculture: measurement and modelling*, M.K. Shukla, Ed., CABI, Wallingford, UK,
19 384–414.
- 20 Fay, P. A., D. M. Kaufman, J. B. Nippert, J. D. Carlisle, and C. W. Harper, 2008: Changes in grassland
21 ecosystem function due to extreme rainfall events: Implications for responses to climate change. *Glob.*
22 *Chang. Biol.*, **14**, 1600–1608, doi:10.1111/j.1365-2486.2008.01605.x.
23 <http://doi.wiley.com/10.1111/j.1365-2486.2008.01605.x> (Accessed May 25, 2018).
- 24 Feddema, J. J., K. W. Oleson, G. B. Bonan, L. O. Mearns, L. E. Buja, G. a Meehl, and W. M. Washington,
25 2005: The importance of land-cover change in simulating future climates. *Science*, **310**, 1674–1678,
26 doi:10.1126/science.1118160.
- 27 Federici, S., F. N. Tubiello, M. Salvatore, H. Jacobs, and J. Schmidhuber, 2015: New estimates of CO₂
28 forest emissions and removals: 1990–2015. *For. Ecol. Manage.*, **352**, 89–98,
29 doi:10.1016/j.foreco.2015.04.022.
- 30 Feng, W., A. F. Plante, and J. Six, 2013: Improving estimates of maximal organic carbon stabilization by
31 fine soil particles. *Biogeochemistry*, **112**, 81–93, doi:10.1007/s10533-011-9679-7.
- 32 Feng, Z. Z., and Coauthors, 2015: Constraints to nitrogen acquisition of terrestrial plants under elevated
33 CO₂. *Glob. Chang. Biol.*, **21**, 3152–3168, doi:10.1111/gcb.12938.
- 34 Fensholt, R., and Coauthors, 2012: Greenness in semi-arid areas across the globe 1981–2007 - an Earth
35 Observing Satellite based analysis of trends and drivers. *Remote Sens. Environ.*, **121**, 144–158,
36 doi:10.1016/j.rse.2012.01.017.
37 <https://www.sciencedirect.com/science/article/abs/pii/S0034425712000545> (Accessed October 31,
38 2018).
- 39 Fernandes, K., L. Verchot, W. Baethgen, V. Gutierrez-Velez, M. Pinedo-Vasquez, and C. Martius, 2017:
40 Heightened fire probability in Indonesia in non-drought conditions: The effect of increasing
41 temperatures. *Environ. Res. Lett.*, **12**, doi:10.1088/1748-9326/aa6884.
- 42 Feyisa, G. L., K. Dons, and H. Meilby, 2014: Efficiency of parks in mitigating urban heat island effect: An
43 example from Addis Ababa. *Landsc. Urban Plan.*, **123**, 87–95,
44 doi:<https://doi.org/10.1016/j.landurbplan.2013.12.008>.
45 <http://www.sciencedirect.com/science/article/pii/S0169204613002399>.
- 46 Ficklin, D. L., and K. A. Novick, 2017: Historic and projected changes in vapor pressure deficit suggest a
47 continental-scale drying of the United States atmosphere. *J. Geophys. Res.*, **122**, 2061–2079,
48 doi:10.1002/2016JD025855. <http://doi.wiley.com/10.1002/2016JD025855> (Accessed October 29,
49 2018).
- 50 Field, C. B., R. B. Jackson, and H. A. Mooney, 1995: Stomatal responses to increased CO₂: implications
51 from the plant to the global scale. *Plant. Cell Environ.*, **18**, 1214–1225, doi:10.1111/j.1365-
52 3040.1995.tb00630.x.
- 53 Findell, K. L., E. Shevliakova, P. C. D. Milly, and R. J. Stouffer, 2007: Modeled impact of anthropogenic
54 land cover change on climate. *J. Clim.*, **20**, 3621–3634, doi:10.1175/JCLI4185.1.
- 55 Findell, K. L., A. J. Pitman, M. H. England, and P. J. Pegion, 2009: Regional and global impacts of land
56 cover change and sea surface temperature anomalies. *J. Clim.*, **22**, 3248–3269,

- 1 doi:10.1175/2008JCLI2580.1.
- 2 Findell, K. L., A. Berg, P. Gentine, J. P. Krasting, B. R. Lintner, S. Malyshev, J. A. Santanello, and E.
3 Shevliakova, 2017: The impact of anthropogenic land use and land cover change on regional climate
4 extremes. *Nat. Commun.*, **8**, 989, doi:10.1038/s41467-017-01038-w.
- 5 Finkbeiner, M., 2014: Indirect land use change – Help beyond the hype? *Biomass and Bioenergy*, **62**, 218–
6 221, doi:http://dx.doi.org/10.1016/j.biombioe.2014.01.024.
7 http://www.sciencedirect.com/science/article/pii/S0961953414000257.
- 8 Finzi, A. C., R. Z. Abramoff, K. S. Spiller, E. R. Brzostek, B. A. Darby, M. A. Kramer, and R. P. Phillips,
9 2015: Rhizosphere processes are quantitatively important components of terrestrial carbon and nutrient
10 cycles. *Glob. Chang. Biol.*, **21**, 2082–2094, doi:10.1111/gcb.12816. https://doi.org/10.1111/gcb.12816.
- 11 Fiore, A. M., and Coauthors, 2012: Global air quality and climate. *Chem. Soc. Rev.*, **41**, 6663,
12 doi:10.1039/c2cs35095e. http://xlink.rsc.org/?DOI=c2cs35095e.
- 13 Fischer, E. M., and R. Knutti, 2014: Detection of spatially aggregated changes in temperature and
14 precipitation extremes. *Geophys. Res. Lett.*, **41**, 547–554, doi:10.1002/2013GL058499.
15 http://doi.wiley.com/10.1002/2013GL058499 (Accessed May 21, 2018).
- 16 —, and —, 2015: Anthropogenic contribution to global occurrence of heavy-precipitation and high-
17 temperature extremes. *Nat. Clim. Chang.*, **5**, 560–564, doi:10.1038/nclimate2617.
- 18 —, J. Rajczak, and C. Schär, 2012: Changes in European summer temperature variability revisited.
19 *Geophys. Res. Lett.*, **39**, doi:10.1029/2012GL052730.
- 20 —, U. Beyerle, and R. Knutti, 2013: Robust spatially aggregated projections of climate extremes. *Nat.*
21 *Clim. Chang.*, **3**, 1033–1038, doi:10.1038/nclimate2051.
- 22 Flanner, M. G., K. M. Shell, M. Barlage, D. K. Perovich, and M. A. Tschudi, 2011: Radiative forcing and
23 albedo feedback from the Northern Hemisphere cryosphere between 1979 and 2008. *Nat. Geosci.*, **4**,
24 151–155, doi:10.1038/ngeo1062. http://dx.doi.org/10.1038/ngeo1062.
- 25 Fletcher, M.-S., S. W. Wood, and S. G. Haberle, 2014: A fire-driven shift from forest to non-forest: evidence
26 for alternative stable states? *Ecology*, **95**, 2504–2513, doi:10.1890/12-1766.1.
27 https://doi.org/10.1890/12-1766.1.
- 28 Forsell, N., O. Turkovska, M. Gusti, M. Obersteiner, M. Den Elzen, and P. Havlik, 2016: Assessing the
29 INDCs' land use, land use change, and forest emission projections. *Carbon Balance Manag.*, **11**,
30 doi:10.1186/s13021-016-0068-3.
- 31 Forzieri, G., R. Alkama, D. G. Miralles, and A. Cescatti, 2017: Satellites reveal contrasting responses of
32 regional climate to the widespread greening of Earth. *Science (80-.)*, **356**, 1180–1184,
33 doi:10.1126/science.aal1727.
- 34 Founda, D., and M. Santamouris, 2017: Synergies between Urban Heat Island and Heat Waves in Athens
35 (Greece), during an extremely hot summer (2012). *Sci. Rep.*, **7**, 1–11, doi:10.1038/s41598-017-11407-
36 6. http://dx.doi.org/10.1038/s41598-017-11407-6.
- 37 Founda, D., F. Pierros, M. Petrakis, and C. Zerefos, 2015: Interdecadal variations and trends of the Urban
38 Heat Island in Athens (Greece) and its response to heat waves. *Atmos. Res.*, **161–162**, 1–13,
39 doi:10.1016/j.atmosres.2015.03.016.
- 40 Fowler, D., and Coauthors, 2015: Effects of global change during the 21st century on the nitrogen cycle.
41 *Atmos. Chem. Phys.*, **15**, 13849–13893, doi:10.5194/acp-15-13849-2015. https://www.atmos-chem-
42 phys.net/15/13849/2015/.
- 43 Frank, D., and Coauthors, 2015a: Effects of climate extremes on the terrestrial carbon cycle: Concepts,
44 processes and potential future impacts. *Glob. Chang. Biol.*, **21**, 2861–2880, doi:10.1111/gcb.12916.
45 http://doi.wiley.com/10.1111/gcb.12916 (Accessed October 16, 2018).
- 46 —, and Coauthors, 2015b: Effects of climate extremes on the terrestrial carbon cycle: Concepts, processes
47 and potential future impacts. *Glob. Chang. Biol.*, **21**, 2861–2880, doi:10.1111/gcb.12916.
- 48 Frank, S., and Coauthors, 2017a: Reducing greenhouse gas emissions in agriculture without compromising
49 food security? *Environ. Res. Lett.*, **12**, 105004, doi:10.1088/1748-9326/aa8c83.
50 http://stacks.iop.org/1748-9326/12/i=10/a=105004?key=crossref.000b6ab1748d1b0af07f66e0f496f0a3
51 (Accessed November 4, 2018).
- 52 —, and Coauthors, 2017b: Reducing greenhouse gas emissions in agriculture without compromising food
53 security? *Environ. Res. Lett.*, **12**, 105004, doi:10.1088/1748-9326/aa8c83.
- 54 Frey, S. D., J. Lee, J. M. Melillo, and J. Six, 2013: The temperature response of soil microbial efficiency and
55 its feedback to climate. *Nat. Clim. Chang.*, **3**, 395, doi:10.1038/nclimate1796.
56 http://www.nature.com/doi/10.1038/nclimate1796.

- 1 Fricko, O., and Coauthors, 2017: The marker quantification of the Shared Socioeconomic Pathway 2: A
2 middle-of-the-road scenario for the 21st century. *Glob. Environ. Chang.*, **42**, 251–267,
3 doi:10.1016/J.GLOENVCHA.2016.06.004.
- 4 Friedlingstein, P., M. Meinshausen, V. K. Arora, C. D. Jones, A. Anav, S. K. Liddicoat, and R. Knutti, 2014:
5 Uncertainties in CMIP5 climate projections due to carbon cycle feedbacks. *J. Clim.*, doi:10.1175/JCLI-
6 D-12-00579.1.
- 7 FRIEDLINGSTEIN, P., and Coauthors, 2006: Climate-carbon cycle feedback analysis: results from the
8 C4MIP model intercomparison. *J. Clim.*, **19**, 3337–3353.
- 9 Friend, A. D., and Coauthors, 2014a: Carbon residence time dominates uncertainty in terrestrial vegetation
10 responses to future climate and atmospheric CO₂. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 3280–3285,
11 doi:10.1073/pnas.1222477110.
- 12 Friend, A. D., and Coauthors, 2014b: Carbon residence time dominates uncertainty in terrestrial vegetation
13 responses to future climate and atmospheric CO₂. *Proc. Natl. Acad. Sci.*, **111**, 3280–3285,
14 doi:10.1073/pnas.1222477110. <http://www.pnas.org/lookup/doi/10.1073/pnas.1222477110>.
- 15 Fritz, S., and Coauthors, 2013: Downgrading recent estimates of land available for biofuel production.
16 *Environ. Sci. Technol.*, doi:10.1021/es303141h.
- 17 Fröhlich-Nowoisky, J., and Coauthors, 2016: Bioaerosols in the Earth system: Climate, health, and
18 ecosystem interactions. *Atmos. Res.*, **182**, 346–376, doi:10.1016/j.atmosres.2016.07.018.
- 19 Froidevaux, P., and Coauthors, 2014: Influence of the Background Wind on the Local Soil Moisture–
20 Precipitation Feedback. *J. Atmos. Sci.*, **71**, 782–799, doi:10.1175/JAS-D-13-0180.1.
- 21 Fu, R., and Coauthors, 2013: Increased dry-season length over southern Amazonia in recent decades and its
22 implication for future climate projection. *Proc. Natl. Acad. Sci.*, **110**, 18110–18115,
23 doi:10.1073/pnas.1302584110.
- 24 Fu, Y., and H. Liao, 2014: Impacts of land use and land cover changes on biogenic emissions of volatile
25 organic compounds in China from the late 1980s to the mid-2000s: implications for tropospheric ozone
26 and secondary organic aerosol. *Tellus B Chem. Phys. Meteorol.*, **66**, 24987,
27 doi:10.3402/tellusb.v66.24987. <https://doi.org/10.3402/tellusb.v66.24987>.
- 28 Fuentes, J. D., and Coauthors, 2000: Biogenic Hydrocarbons in the Atmospheric Boundary Layer: A
29 Review. *Bull. Am. Meteorol. Soc.*, **81**, 1537–1575, doi:10.1175/1520-
30 0477(2000)081<1537:BHITAB>2.3.CO;2. [https://doi.org/10.1175/1520-
31 0477\(2000\)081%3C1537:BHITAB%3E2.3.CO](https://doi.org/10.1175/1520-0477(2000)081%3C1537:BHITAB%3E2.3.CO).
- 32 Fuentes, J. D., and Coauthors, 2016: Linking meteorology, turbulence, and air chemistry in the amazon rain
33 forest. *Bull. Am. Meteorol. Soc.*, **97**, 2329–2342, doi:10.1175/BAMS-D-15-00152.1.
- 34 Fuglestedt, J., and Coauthors, 2018: Implications of possible interpretations of “greenhouse gas balance” in
35 the Paris Agreement. *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, **376**, 20160445,
36 doi:10.1098/rsta.2016.0445.
- 37 Fujibe, F., 2009: Detection of urban warming in recent temperature trends in Japan. *Int. J. Climatol.*, **29**,
38 1811–1822, doi:10.1002/joc.1822. <https://doi.org/10.1002/joc.1822>.
- 39 Fujimori, S., T. Hasegawa, T. Masui, K. Takahashi, D. S. Herran, H. Dai, Y. Hijioka, and M. Kainuma,
40 2017: SSP3: AIM implementation of Shared Socioeconomic Pathways. *Glob. Environ. Chang.*, **42**,
41 268–283, doi:10.1016/J.GLOENVCHA.2016.06.009.
- 42 Funk, C., and Coauthors, 2018: Anthropogenic Enhancement of Moderate-to-Strong El Niño Events Likely
43 Contributed to Drought and Poor Harvests in Southern Africa During 2016. *Bull. Am. Meteorol. Soc.*,
44 **99**, S91–S96.
- 45 Fuss, S., and Coauthors, 2018: Negative emissions - Part 2: Costs, potentials and side effects. *Environ. Res.*
46 *Lett.*, **submitted**, 2–4, doi:<https://doi.org/10.1088/1748-9326/aabf9f>. [http://stacks.iop.org/1748-
47 9326/13/i=6/a=063002](http://stacks.iop.org/1748-9326/13/i=6/a=063002).
- 48 Fuzzi, S., and Coauthors, 2015a: Particulate matter, air quality and climate: Lessons learned and future
49 needs. *Atmos. Chem. Phys.*, **15**, 8217–8299, doi:10.5194/acp-15-8217-2015.
- 50 ———, and Coauthors, 2015b: Particulate matter, air quality and climate: Lessons learned and future needs.
51 *Atmos. Chem. Phys.*, **15**, 8217–8299, doi:10.5194/acp-15-8217-2015. [https://www.atmos-chem-
52 phys.net/15/8217/2015/](https://www.atmos-chem-phys.net/15/8217/2015/).
- 53 Gallego-Sala, A. V., and Coauthors, 2018: Latitudinal limits to the predicted increase of the peatland carbon
54 sink with warming. *Nat. Clim. Chang.*, **8**, 907–913, doi:10.1038/s41558-018-0271-1.
- 55 Gálos, B., C. Mátyás, and D. Jacob, 2011: Regional characteristics of climate change altering effects of
56 afforestation. *Environ. Res. Lett.*, **6**, 44010, doi:10.1088/1748-9326/6/4/044010.

- 1 ———, and Coauthors, 2013: Case study for the assessment of the biogeophysical effects of a potential
2 afforestation in Europe. *Carbon Balance Manag.*, **8**, 3, doi:10.1186/1750-0680-8-3.
- 3 Ganeshan, M., and R. Murtugudde, 2015: Nocturnal propagating thunderstorms may favor urban “hot-
4 spots”: A model-based study over Minneapolis. *Urban Clim.*, **14**, 606–621,
5 doi:10.1016/j.uclim.2015.10.005. <http://dx.doi.org/10.1016/j.uclim.2015.10.005>.
- 6 Ganopolski, A., V. Petoukhov, S. Rahmstorf, V. Brovkin, M. Claussen, A. Eliseev, and C. Kubatzki, 2001a:
7 Climber-2: A climate system model of intermediate complexity. Part II: Model sensitivity. *Clim. Dyn.*,
8 **17**, 735–751, doi:10.1007/s003820000144.
- 9 ———, ———, ———, ———, ———, ———, and ———, 2001b: Climber-2: A climate system model of intermediate
10 complexity. Part II: Model sensitivity. *Clim. Dyn.*, **17**, 735–751, doi:10.1007/s003820000144.
- 11 ———, ———, ———, ———, ———, a. Eliseev, and C. Kubatzki, 2001c: Climber-2: A climate system model of
12 intermediate complexity. Part II: Model sensitivity. *Clim. Dyn.*, **17**, 735–751,
13 doi:10.1007/s003820000144.
- 14 Gao, Y., T. Markkanen, L. Backman, H. M. Henttonen, J. P. Pietikäinen, H. M. Mäkelä, and A. Laaksonen,
15 2014: Biogeophysical impacts of peatland forestation on regional climate changes in Finland.
16 *Biogeosciences*, **11**, 7251–7267, doi:10.5194/bg-11-7251-2014.
- 17 García-Palacios, P., and Coauthors, 2015: Are there links between responses of soil microbes and ecosystem
18 functioning to elevated CO₂, N deposition and warming? A global perspective. *Glob. Chang. Biol.*, **21**,
19 1590–1600, doi:10.1111/gcb.12788.
- 20 García-Ruiz, J. M., I. I. López-Moreno, S. M. Vicente-Serrano, T. Lasanta-Martínez, and S. Beguería, 2011:
21 Mediterranean water resources in a global change scenario. *Earth-Science Rev.*, **105**, 121–139,
22 doi:10.1016/j.earscirev.2011.01.006.
23 <https://www.sciencedirect.com/science/article/pii/S0012825211000134> (Accessed September 24,
24 2018).
- 25 Garcia, R. A., M. Cabeza, R. Altwegg, and M. B. Araújo, 2016: Do projections from bioclimatic envelope
26 models and climate change metrics match? *Glob. Ecol. Biogeogr.*, **25**, 65–74, doi:10.1111/geb.12386.
27 <http://doi.wiley.com/10.1111/geb.12386> (Accessed April 6, 2019).
- 28 Garnaud, C., and L. Sushama, 2015: Biosphere-climate interactions in a changing climate over North
29 America. *J. Geophys. Res. Atmos.*, **120**, 1091–1108, doi:10.1002/2014JD022055.
- 30 Gasser, T., and P. Ciais, 2013: A theoretical framework for the net land-to-atmosphere CO₂ flux and its
31 implications in the definition of “emissions from land-use change.” *Earth Syst. Dyn.*, **4**, 171–186,
32 doi:10.5194/esd-4-171-2013. <https://www.earth-syst-dynam.net/4/171/2013/>.
- 33 Gatti, L. V., and Coauthors, 2014: Drought sensitivity of Amazonian carbon balance revealed by
34 atmospheric measurements. *Nature*, doi:10.1038/nature12957.
- 35 Gauthier, S., P. Bernier, T. Kuuluvainen, A. Z. Shvidenko, and D. G. Schepaschenko, 2015: Boreal forest
36 health and global change. *Science (80-.)*, **349**, 819–822, doi:10.1126/science.aaa9092.
- 37 Gedalof, Z., and A. A. Berg, 2010: Tree ring evidence for limited direct CO₂ fertilization of forests over the
38 20th century. *Global Biogeochem. Cycles*, **24**, n/a-n/a, doi:10.1029/2009GB003699.
39 <http://doi.wiley.com/10.1029/2009GB003699>.
- 40 Geddes, A., W. N. Adger, N. W. Arnell, R. Black, and D. S. G. Thomas, 2012: Migration, Environmental
41 Change, and the “Challenges of Governance.” *Environ. Plan. C Gov. Policy*, **30**, 951–967,
42 doi:10.1068/c3006ed. <https://doi.org/10.1068/c3006ed>.
- 43 Gedney, N., and P. J. Valdes, 2000: The effect of Amazonian deforestation on the northern hemisphere
44 circulation and climate. *Geophys. Res. Lett.*, **27**, 3053, doi:10.1029/2000GL011794.
- 45 Geirinhas, J. L., R. M. Trigo, R. Libonati, C. A. S. Coelho, and A. C. Palmeira, 2018: Climatic and synoptic
46 characterization of heat waves in Brazil. *Int. J. Climatol.*, **38**, 1760–1776, doi:10.1002/joc.5294.
- 47 Gellesch, E., M. A. S. Arfin Khan, J. Kreyling, A. Jentsch, and C. Beierkuhnlein, 2017: Grassland
48 experiments under climatic extremes: Reproductive fitness versus biomass. *Environ. Exp. Bot.*, **144**,
49 68–75, doi:10.1016/j.envexpbot.2017.10.007.
50 <https://www.sciencedirect.com/science/article/abs/pii/S0098847217302447> (Accessed October 31,
51 2018).
- 52 Georgescu, M., D. B. Lobell, and C. B. Field, 2011: Direct climate effects of perennial bioenergy crops in
53 the United States. *Proc. Natl. Acad. Sci.*, **108**, 4307–4312, doi:10.1073/pnas.1008779108.
54 <http://www.pnas.org/content/early/2011/02/16/1008779108.abstract>.
- 55 Georgescu, M., D. B. Lobell, C. B. Field, and A. Mahalov, 2013: Simulated hydroclimatic impacts of
56 projected Brazilian sugarcane expansion. *Geophys. Res. Lett.*, **40**, 972–977, doi:10.1002/grl.50206.

- 1 Gerbrandt, K., P. L. Chu, A. Simmonds, K. A. Mullins, H. L. MacLean, W. M. Griffin, and B. A. Saville,
2 2016: Life cycle assessment of lignocellulosic ethanol: a review of key factors and methods affecting
3 calculated GHG emissions and energy use. *Curr. Opin. Biotechnol.*, **38**, 63–70,
4 doi:<https://doi.org/10.1016/j.copbio.2015.12.021>.
5 <http://www.sciencedirect.com/science/article/pii/S0958166916000033>.
- 6 Gerster-Bentaya, M., 2013: Nutrition-sensitive urban agriculture. *Food Secur.*, **5**, 723–737,
7 doi:10.1007/s12571-013-0295-3. <https://doi.org/10.1007/s12571-013-0295-3>.
- 8 Gestel, N. Van, and Coauthors, 2018: Brief Communications Arising Predicting soil carbon loss with
9 warming. *Nat. Publ. Gr.*, **554**, E4–E5, doi:10.1038/nature25745.
10 <http://dx.doi.org/10.1038/nature25745>.
- 11 Gharun, M., M. Possell, T. L. Bell, and M. A. Adams, 2017: Optimisation of fuel reduction burning regimes
12 for carbon, water and vegetation outcomes. *J. Environ. Manage.*, **203**, 157–170,
13 doi:<https://doi.org/10.1016/j.jenvman.2017.07.056>.
- 14 Gibbard, S., K. Caldeira, G. Bala, T. J. Phillips, and M. Wickett, 2005: Climate effects of global land cover
15 change. *Geophys. Res. Lett.*, **32**, 1–4, doi:10.1029/2005GL024550.
- 16 Gidden, M. J., and Coauthors, 2018: Global emissions pathways under different socioeconomic scenarios for
17 use in CMIP6: a dataset of harmonized emissions trajectories through the end of the century. *Geosci.*
18 *Model Dev. Discuss.*, 1–42, doi:10.5194/gmd-2018-266.
- 19 Giglio, L., J. T. Randerson, and G. R. Van Der Werf, 2013: Analysis of daily, monthly, and annual burned
20 area using the fourth-generation global fire emissions database (GFED4). *J. Geophys. Res.*
21 *Biogeosciences*, **118**, 317–328, doi:10.1002/jgrg.20042.
- 22 Gill, S. E., J. F. Handley, A. R. Ennos, and S. Pauleit, 2007: Adapting cities for climate change: The role of
23 the green infrastructure. *Built Environ.*, **33**, 115–133, doi:10.2148/benv.33.1.115.
- 24 Gillett, N. P., V. K. Arora, D. Matthews, and M. R. Allen, 2013: Constraining the ratio of global warming to
25 cumulative CO₂ emissions using CMIP5 simulations. *J. Clim.*, **26**, 6844–6858, doi:10.1175/JCLI-D-
26 12-00476.1. <http://dx.doi.org/10.1175/JCLI-D-12-00476.1>.
- 27 Ginoux, P., J. M. Prospero, T. E. Gill, N. C. Hsu, and M. Zhao, 2012: Global-scale attribution of
28 anthropogenic and natural dust sources and their emission rates based on MODIS Deep Blue aerosol
29 products. *Rev. Geophys.*, **50**, doi:10.1029/2012RG000388.
- 30 Goetz, S. J., M. C. Hansen, R. A. Houghton, W. Walker, N. Laporte, and J. Busch, 2015: Measurement and
31 monitoring needs , capabilities and potential for addressing reduced emissions from deforestation and
32 forest degradation under Measurement and monitoring needs , capabilities and potential for addressing
33 reduced emissions from deforestation. *Environ. Res. Lett.*, **10**.
- 34 Goldstein, B., M. Birkved, J. Fernández, and M. Hauschild, 2017: Surveying the Environmental Footprint of
35 Urban Food Consumption. *J. Ind. Ecol.*, **21**, 151–165, doi:10.1111/jiec.12384.
- 36 Gonçalves, K. dos S. K. D. S. D. S., M. S. M. S. S. M. S. Winkler, P. R. P. R. R. Benchimol-Barbosa, K. de
37 Hoogh, P. E. P. E. E. Artaxo, S. de Souza Hacon, C. Schindler, and N. Künzli, 2018: Development of
38 non-linear models predicting daily fine particle concentrations using aerosol optical depth retrievals
39 and ground-based measurements at a municipality in the Brazilian Amazon region. *Atmos. Environ.*,
40 **184**, 156–165, doi:10.1016/j.atmosenv.2018.03.057.
- 41 Gonsamo, A., J. M. Chen, Y. W. Ooi, C. J. M., and O. Y. W., 2017: Peak season plant activity shift towards
42 spring is reflected by increasing carbon uptake by extratropical ecosystems. *Glob. Chang. Biol.*, **24**,
43 2117–2128, doi:10.1111/gcb.14001.
- 44 González-Jaramillo, V., A. Fries, R. Rollenbeck, J. Paladines, F. Oñate-Valdivieso, and J. Bendix, 2016:
45 Assessment of deforestation during the last decades in Ecuador using NOAA-AVHRR satellite data.
46 *Erdkunde*, **70**, 217–235, doi:10.3112/erdkunde.2016.03.02.
- 47 Gottfried, M., and Coauthors, 2012: Continent-wide response of mountain vegetation to climate change. *Nat.*
48 *Clim. Chang.*, **2**, 111–115, doi:10.1038/nclimate1329. <http://www.nature.com/articles/nclimate1329>
49 (Accessed March 12, 2019).
- 50 Govaerts, B., N. Verhulst, A. Castellanos-Navarrete, K. D. Sayre, J. Dixon, and L. Dendooven, 2009:
51 Conservation agriculture and soil carbon sequestration: Between myth and farmer reality. *CRC. Crit.*
52 *Rev. Plant Sci.*, **28**, 97–122, doi:10.1080/07352680902776358.
- 53 Grant, K., J. Kreyling, H. Heilmeyer, C. Beierkuhnlein, and A. Jentsch, 2014: Extreme weather events and
54 plant–plant interactions: shifts between competition and facilitation among grassland species in the face
55 of drought and heavy rainfall. *Ecol. Res.*, **29**, 991–1001, doi:10.1007/s11284-014-1187-5.
56 <http://link.springer.com/10.1007/s11284-014-1187-5> (Accessed October 31, 2018).

- 1 Grassi, G., M. G. J. den Elzen, A. F. Hof, R. Pilli, and S. Federici, 2012: The role of the land use, land use
2 change and forestry sector in achieving Annex I reduction pledges. *Clim. Change*, **115**, 873–881,
3 doi:10.1007/s10584-012-0584-4. <https://doi.org/10.1007/s10584-012-0584-4>.
- 4 ———, J. House, F. Dentener, S. Federici, M. Den Elzen, and J. Penman, 2017: The key role of forests in
5 meeting climate targets requires science for credible mitigation. *Nat. Clim. Chang.*, **7**, 220–226,
6 doi:10.1038/nclimate3227. <http://www.nature.com/doi/10.1038/nclimate3227>.
- 7 ———, ———, W. A. Kurz, A. Cescatti, and R. A. Houghton, 2018a: Reconciling global model estimates and
8 country reporting of anthropogenic forest CO₂ sinks. *Nat. Clim. Chang.*, 1–35.
- 9 ———, R. Pilli, J. House, S. Federici, and W. A. Kurz, 2018b: Science-based approach for credible accounting
10 of mitigation in managed forests. *Carbon Balance Manag.*, doi:10.1186/s13021-018-0096-2.
- 11 Gray, S. B., and Coauthors, 2016: Intensifying drought eliminates the expected benefits of elevated carbon
12 dioxide for soybean. *Nat. Plants*, **2**, 16132, doi:10.1038/nplants.2016.132.
13 <http://www.nature.com/articles/nplants2016132> (Accessed May 10, 2018).
- 14 Green, J. K., S. I. Seneviratne, A. M. Berg, K. L. Findell, S. Hagemann, D. M. Lawrence, and P. Gentine,
15 2019: Large influence of soil moisture on long-term terrestrial carbon uptake. *Nature*, **565**, 476–479,
16 doi:10.1038/s41586-018-0848-x.
- 17 Gregory, P. J., and B. Marshall, 2012: Attribution of climate change: A methodology to estimate the
18 potential contribution to increases in potato yield in Scotland since 1960. *Glob. Chang. Biol.*, **18**, 1372–
19 1388, doi:10.1111/j.1365-2486.2011.02601.x. <http://doi.wiley.com/10.1111/j.1365-2486.2011.02601.x>
20 (Accessed May 10, 2018).
- 21 Gren, I.-M., and A. Z. Aklilu, 2016: Policy design for forest carbon sequestration: A review of the literature.
22 *For. Policy Econ.*, **70**, 128–136, doi:<https://doi.org/10.1016/j.forpol.2016.06.008>.
23 <http://www.sciencedirect.com/science/article/pii/S1389934116301253>.
- 24 Grillakis, M. G., 2019: Increase in severe and extreme soil moisture droughts for Europe under climate
25 change. *Sci. Total Environ.*, **660**, 1245–1255, doi:10.1016/j.scitotenv.2019.01.001.
- 26 Griscom, B. W., and Coauthors, 2017a: Natural climate solutions. *Proc. Natl. Acad. Sci.*, **114**, 11645–11650,
27 doi:10.1073/pnas.1710465114.
- 28 ———, and Coauthors, 2017b: Natural climate solutions. *Proc. Natl. Acad. Sci.*, **114**, 11645–11650,
29 doi:10.1073/pnas.1710465114. <http://www.pnas.org/lookup/doi/10.1073/pnas.1710465114>.
- 30 ———, and Coauthors, 2017c: Natural climate solutions. *Proc. Natl. Acad. Sci.*, **114**, 11645–11650,
31 doi:10.1073/pnas.1710465114.
- 32 van Groenigen, K. J., X. Qi, C. W. Osenberg, Y. Luo, and B. A. Hungate, 2014: Faster Decomposition
33 Under Increased Atmospheric CO₂ Limits Soil Carbon Storage. *Science (80-.)*, **344**, 508–509,
34 doi:10.1126/science.1249534.
- 35 Groß, S., V. Freudenthaler, K. Schepanski, C. Toledano, A. Schäfler, A. Ansmann, and B. Weinzierl, 2015:
36 Optical properties of long-range transported Saharan dust over Barbados as measured by dual-
37 wavelength depolarization Raman lidar measurements. *Atmos. Chem. Phys.*, **15**, 11067–11080,
38 doi:10.5194/acp-15-11067-2015.
- 39 Grubler, A., and Coauthors, 2018: A low energy demand scenario for meeting the 1.5 °C target and
40 sustainable development goals without negative emission technologies. *Nat. Energy*, **3**, 515–527,
41 doi:10.1038/s41560-018-0172-6. <http://www.nature.com/articles/s41560-018-0172-6> (Accessed
42 October 15, 2018).
- 43 Guan, K., B. Sultan, M. Biasutti, C. Baron, and D. B. Lobell, 2015: What aspects of future rainfall changes
44 matter for crop yields in West Africa? *Geophys. Res. Lett.*, **42**, 8001–8010,
45 doi:10.1002/2015GL063877.
- 46 Gudipudi, R., D. Rybski, M. K. B. Lüdeke, B. Zhou, Z. Liu, and J. P. Kropp, 2019: The efficient, the
47 intensive, and the productive: Insights from urban Kaya scaling. *Appl. Energy*, **236**, 155–162,
48 doi:10.1016/j.apenergy.2018.11.054.
- 49 Guenet, B., M. Camino-Serrano, P. Ciais, M. Tifafi, F. Maignan, J. L. Soong, and I. A. Janssens, 2018:
50 Impact of priming on global soil carbon stocks. *Glob. Chang. Biol.*, **24**, 1873–1883,
51 doi:10.1111/gcb.14069. <https://doi.org/10.1111/gcb.14069>.
- 52 Guenther, A., and Coauthors, 1995: A global model of natural volatile organic compound emissions. *J.*
53 *Geophys. Res.*, **100**, 8873–8892, doi:10.1029/94JD02950. <https://doi.org/10.1029/94JD02950>.
- 54 Guenther, A. B., X. Jiang, C. L. Heald, T. Sakulyanontvittaya, T. Duhl, L. K. Emmons, and X. Wang, 2012:
55 The model of emissions of gases and aerosols from nature version 2.1 (MEGAN2.1): An extended and
56 updated framework for modeling biogenic emissions. *Geosci. Model Dev.*, **5**, 1471–1492,

- 1 doi:10.5194/gmd-5-1471-2012.
- 2 Guest, G., F. Cherubini, and A. H. Strømman, 2013: The role of forest residues in the accounting for the
3 global warming potential of bioenergy. *GCB Bioenergy*, **5**, 459–466, doi:doi: 10.1111/gcbb.12014.
- 4 Guhathakurta, P., D. S. Pai, and M. N. Rajeevan, 2017: Variability and Trends of Extreme Rainfall and
5 Rainstorms. Springer, Singapore, 37–49 http://link.springer.com/10.1007/978-981-10-2531-0_3
6 (Accessed May 24, 2018).
- 7 Guillod, B. P., B. Orlowsky, D. G. Miralles, A. J. Teuling, and S. I. Seneviratne, 2015: Reconciling spatial
8 and temporal soil moisture effects on afternoon rainfall. *Nat. Commun.*, **6**, 6443,
9 doi:10.1038/ncomms7443.
- 10 Guimberteau, M., K. Laval, A. Perrier, and J. Polcher, 2012: Global effect of irrigation and its impact on the
11 onset of the Indian summer monsoon. *Clim. Dyn.*, **39**, 1329–1348, doi:10.1007/s00382-011-1252-5.
- 12 Gumbrecht, T., R. M. Roman-Cuesta, L. Verchot, M. Herold, F. Wittmann, E. Householder, N. Herold, and
13 D. Murdiyarso, 2017: An expert system model for mapping tropical wetlands and peatlands reveals
14 South America as the largest contributor. *Glob. Chang. Biol.*, **23**, 3581–3599, doi:10.1111/gcb.13689.
15 <https://doi.org/10.1111/gcb.13689>.
- 16 Guo, Z., and Coauthors, 2006: GLACE: The Global Land–Atmosphere Coupling Experiment. Part II:
17 Analysis. *J. Hydrometeorol.*, **7**, 611–625, doi:10.1175/JHM511.1.
- 18 Gustafsson, Ö., and V. Ramanathan, 2016: Convergence on climate warming by black carbon aerosols. *Proc.*
19 *Natl. Acad. Sci.*, **113**, 4243–4245, doi:10.1073/pnas.1603570113.
20 <http://www.pnas.org/lookup/doi/10.1073/pnas.1603570113>.
- 21 Gustavsson, L., S. Haus, M. Lundblad, A. Lundström, C. A. Ortiz, R. Sathre, N. Le Truong, and P. E.
22 Wikberg, 2017: Climate change effects of forestry and substitution of carbon-intensive materials and
23 fossil fuels. *Renew. Sustain. Energy Rev.*, doi:10.1016/j.rser.2016.09.056.
- 24 Gütschow, J., M. L. Jeffery, R. Gieseke, R. Gebel, D. Stevens, M. Krapp, and M. Rocha, 2016: The
25 PRIMAP-hist national historical emissions time series. *Earth Syst. Sci. Data*, **8**, 571–603,
26 doi:10.5194/essd-8-571-2016.
- 27 Haberl, H., 2013: Net land-atmosphere flows of biogenic carbon related to bioenergy: Towards an
28 understanding of systemic feedbacks. *GCB Bioenergy*.
- 29 ———, and Coauthors, 2012: Correcting a fundamental error in greenhouse gas accounting related to
30 bioenergy. *Energy Policy*, doi:10.1016/j.enpol.2012.02.051.
- 31 Haddaway, N. R., K. Hedlund, L. E. Jackson, T. Kätterer, E. Lugato, I. K. Thomsen, H. B. Jørgensen, and P.
32 E. Isberg, 2017: How does tillage intensity affect soil organic carbon? A systematic review. *Environ.*
33 *Evid.*, **6**, 30, doi:10.1186/s13750-017-0108-9.
- 34 Hagos, S., L. R. Leung, Y. Xue, A. Boone, F. de Sales, N. Neupane, M. Huang, and J. H. Yoon, 2014:
35 Assessment of uncertainties in the response of the African monsoon precipitation to land use change
36 simulated by a regional model. *Clim. Dyn.*, **43**, 2765–2775, doi:10.1007/s00382-014-2092-x.
- 37 Hamdi, R., 2010: Estimating urban heat island effects on the temperature series of Uccle (Brussels, Belgium)
38 using remote sensing data and a land surface scheme. *Remote Sens.*, **2**, 2773–2784,
39 doi:10.3390/rs2122773.
- 40 ———, P. Termonia, and P. Baguis, 2011: Effects of urbanization and climate change on surface runoff of the
41 Brussels Capital Region: A case study using an urban soil-vegetation-atmosphere-transfer model. *Int. J.*
42 *Climatol.*, **31**, 1959–1974, doi:10.1002/joc.2207.
- 43 Hamdi, R., F. Duchêne, J. Berckmans, A. Delcloo, C. Vanpoucke, and P. Termonia, 2016: Evolution of
44 urban heat wave intensity for the Brussels Capital Region in the ARPEGE-Climat A1B scenario. *Urban*
45 *Clim.*, **17**, 176–195, doi:10.1016/j.uclim.2016.08.001. <http://dx.doi.org/10.1016/j.uclim.2016.08.001>.
- 46 Hamilton, A. J., K. Burry, H.-F. Mok, S. F. Barker, J. R. Grove, and V. G. Williamson, 2014: Give peas a
47 chance? Urban agriculture in developing countries. A review. *Agron. Sustain. Dev.*, **34**, 45–73,
48 doi:10.1007/s13593-013-0155-8. <https://doi.org/10.1007/s13593-013-0155-8>.
- 49 Han, S., and Z. Yang, 2013: Cooling effect of agricultural irrigation over Xinjiang, Northwest China from
50 1959 to 2006. *Environ. Res. Lett.*, **8**, doi:10.1088/1748-9326/8/2/024039.
- 51 Han Weng, Z., and Coauthors, 2017: Biochar built soil carbon over a decade by stabilizing rhizodeposits.
52 *Nat. Clim. Chang.*, **7**, 371–376, doi:10.1038/nclimate3276. <http://dx.doi.org/10.1038/nclimate3276>.
- 53 Hanes, C. C., X. Wang, P. Jain, M.-A. Parisien, J. M. Little, and M. D. Flannigan, 2018: Fire-regime changes
54 in Canada over the last half century. *Can. J. For. Res.*, **49**, 256–269, doi:10.1139/cjfr-2018-0293.
- 55 Hansen, J., and Coauthors, 2005: Efficacy of climate forcings. *J. Geophys. Res. D Atmos.*, **110**, 1–45,
56 doi:10.1029/2005JD005776.

- 1 ———, R. Ruedy, M. Sato, and K. Lo, 2010: Global surface temperature change. *Rev. Geophys.*, **48**, RG4004,
2 doi:10.1029/2010RG000345.
- 3 Hansen, J. E., M. Sato, A. Lacis, R. Ruedy, I. Tegen, and E. Matthews, 1998: Climate forcings in the
4 Industrial era. *Proc. Natl. Acad. Sci.*, **95**, 12753–12758, doi:10.1073/pnas.95.22.12753.
5 <https://www.pnas.org/content/95/22/12753>.
- 6 Hansen, M. C., and Coauthors, 2012: Carbon emissions from land use and land-cover change.
7 *Biogeosciences*, **9**, 5125–5142, doi:10.5194/bg-9-5125-2012.
- 8 Hansis, E., S. J. Davis, and J. Pongratz, 2015: Relevance of methodological choices for accounting of land
9 use change carbon fluxes. *Global Biogeochem. Cycles*, **29**, 1230–1246, doi:10.1002/2014GB004997.
10 <https://agupubs.onlinelibrary.wiley.com/doi/full/10.1002/2014GB004997>.
- 11 Hantson, S., and Coauthors, 2016: The status and challenge of global fire modelling. *Biogeosciences*, **13**,
12 3359–3375, doi:10.5194/bg-13-3359-2016.
- 13 ———, and Coauthors, 2017: Global isoprene and monoterpene emissions under changing climate, vegetation,
14 CO₂ and land use. *Atmos. Environ.*, **155**, 35–45, doi:10.1016/j.atmosenv.2017.02.010.
- 15 Harding, K. J., T. E. Twine, A. VanLoocke, J. E. Bagley, and J. Hill, 2016: Impacts of second-generation
16 biofuel feedstock production in the central U.S. on the hydrologic cycle and global warming mitigation
17 potential. *Geophys. Res. Lett.*, **43**, 10,710–773,781, doi:10.1002/2016gl069981.
18 <http://dx.doi.org/10.1002/2016GL069981>.
- 19 Harlan, S. L., and D. M. Ruddell, 2011: Climate change and health in cities: impacts of heat and air pollution
20 and potential co-benefits from mitigation and adaptation. *Curr. Opin. Environ. Sustain.*, **3**, 126–134,
21 doi:<https://doi.org/10.1016/j.cosust.2011.01.001>.
22 <http://www.sciencedirect.com/science/article/pii/S1877343511000029>.
- 23 Harper, A. B., and Coauthors, 2018: Land-use emissions play a critical role in land-based mitigation for
24 Paris climate targets. *Nature Communications* <https://doi.org/10.1038/s41467-018-05340-z>.
- 25 Harris, R. M. B., T. A. Remenyi, G. J. Williamson, N. L. Bindoff, and D. M. J. S. Bowman, 2016: Climate–
26 vegetation–fire interactions and feedbacks: trivial detail or major barrier to projecting the future of the
27 Earth system? *Wiley Interdiscip. Rev. Clim. Chang.*, **7**, 910–931, doi:10.1002/wcc.428.
28 <https://doi.org/10.1002/wcc.428>.
- 29 Harris, Z. M., R. Spake, and G. Taylor, 2015: Land use change to bioenergy: A meta-analysis of soil carbon
30 and GHG emissions. *Biomass and Bioenergy*, **82**, 27–39, doi:10.1016/j.biombioe.2015.05.008.
31 <http://www.sciencedirect.com/science/article/pii/S0961953415001853>.
- 32 Harrison, A. D., T. F. Whale, M. A. Carpenter, M. A. Holden, L. Neve, D. O’Sullivan, J. Vergara Temprado,
33 and B. J. Murray, 2016: Not all feldspars are equal: a survey of ice nucleating properties across the
34 feldspar group of minerals. *Atmos. Chem. Phys.*, **16**, 10927–10940, doi:10.5194/acp-16-10927-2016.
- 35 Harrison, S. P., J. R. Marlon, and P. J. Bartlein, 2010: Fire in the Earth System. *Chang. Clim. Earth Syst.*
36 *Soc.*, **324**, 21–48, doi:10.1007/978-90-481-8716-4_3. http://link.springer.com/10.1007/978-90-481-8716-4_3.
- 37
- 38 Harsch, M. A., P. E. Hulme, M. S. McGlone, and R. P. Duncan, 2009: Are treelines advancing? A global
39 meta-analysis of treeline response to climate warming. *Ecol. Lett.*, **12**, 1040–1049, doi:10.1111/j.1461-
40 0248.2009.01355.x. <http://doi.wiley.com/10.1111/j.1461-0248.2009.01355.x> (Accessed May 20, 2018).
- 41 Hartley, A. J., D. J. Parker, L. Garcia-Carreras, and S. Webster, 2016: Simulation of vegetation feedbacks on
42 local and regional scale precipitation in West Africa. *Agric. For. Meteorol.*, **222**, 59–70,
43 doi:10.1016/j.agrformet.2016.03.001.
- 44 Hartley, I. P., D. W. Hopkins, M. H. Garnett, M. Sommerkorn, and P. A. Wookey, 2008: Soil microbial
45 respiration in arctic soil does not acclimate to temperature. *Ecol. Lett.*, **11**, 1092–1100,
46 doi:10.1111/j.1461-0248.2008.01223.x.
- 47 ———, M. H. Garnett, M. Sommerkorn, D. W. Hopkins, B. J. Fletcher, V. L. Sloan, G. K. Phoenix, and P. A.
48 Wookey, 2012: A potential loss of carbon associated with greater plant growth in the European Arctic.
49 *Nat. Clim. Chang.*, **2**, 875–879, doi:10.1038/nclimate1575. <http://dx.doi.org/10.1038/nclimate1575>.
- 50 Hartmann, D. L., and Coauthors, 2013a: Observations: Atmosphere and Surface. *Climate Change 2013: The*
51 *Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the*
52 *Intergovernmental Panel on Climate Change*, T.F. Stocker et al., Eds., Cambridge University Pres,
53 Cambridge, United Kingdom and New York, NY, USA, p. 254.
- 54 ———, and Coauthors, 2013b: Observations: Atmosphere and surface. *Climate Change 2013 the Physical*
55 *Science Basis: Working Group I Contribution to the Fifth Assessment Report of the Intergovernmental*
56 *Panel on Climate Change*, T.F. Stocker et al., Eds., Vol. 9781107057 of, Cambridge University Pres,

- 1 Cambridge, United Kingdom and New York, NY, USA, 159–254 <http://www.climatechange2013.org>.
- 2 Hartmann, H., and Coauthors, 2018: Research frontiers for improving our understanding of drought-induced
3 tree and forest mortality. *New Phytol.*, **218**, 15–28, doi:10.1111/nph.15048.
4 <https://doi.org/10.1111/nph.15048>.
- 5 Hassink, J., and A. P. Whitmore, 1997: A Model of the Physical Protection of Organic Matter in Soils. *Soil*
6 *Sci. Soc. Am. J.*, **61**, 131, doi:10.2136/sssaj1997.03615995006100010020x.
7 <https://www.soils.org/publications/sssaj/abstracts/61/1/SS0610010131>.
- 8 Hatfield, J. L., and J. H. Prueger, 2015: Temperature extremes: Effect on plant growth and development.
9 *Weather Clim. Extrem.*, **10**, 4–10, doi:10.1016/J.WACE.2015.08.001.
10 <https://www.sciencedirect.com/science/article/pii/S2212094715300116> (Accessed October 31, 2018).
- 11 Hauser, M., R. Orth, and S. I. Seneviratne, 2017: Investigating soil moisture–climate interactions with
12 prescribed soil moisture experiments: an assessment with the Community Earth System Model (version
13 1.2). *Geosci. Model Dev.*, **10**, 1665–1677, doi:10.5194/gmd-10-1665-2017.
- 14 Hausfather, Z., M. J. Menne, C. N. Williams, T. Masters, R. Broberg, and D. Jones, 2013: Quantifying the
15 effect of urbanization on u.s. Historical climatology network temperature records. *J. Geophys. Res.*
16 *Atmos.*, **118**, 481–494, doi:10.1029/2012JD018509.
- 17 Hausmann, P., R. Sussmann, and D. Smale, 2015: Contribution of oil and natural gas production to renewed
18 increase of atmospheric methane (2007–2014): Top-down estimate from ethane and methane column
19 observations. *Atmos. Chem. Phys. Discuss.*, **15**, 35991–36028, doi:10.5194/acpd-15-35991-2015.
- 20 Hawken, P., 2017: *Drawdown : the most comprehensive plan ever proposed to reverse global warming*.
21 Penguin Books, 240 pp.
- 22 Hawkes, C. V., B. G. Waring, J. D. Rocca, and S. N. Kivlin, 2017: Historical climate controls soil respiration
23 responses to current soil moisture. *Proc. Natl. Acad. Sci.*, **114**, 6322–6327,
24 doi:10.1073/pnas.1620811114. <http://www.pnas.org/content/114/24/6322>.
- 25 Hawkins, E., and Coauthors, 2014: Uncertainties in the timing of unprecedented climates. *Nature*, **511**, E3–
26 E5, doi:10.1038/nature13523.
- 27 He, F., S. J. Vavrus, J. E. Kutzbach, W. F. Ruddiman, J. O. Kaplan, and K. M. Krumhardt, 2014: Simulating
28 global and local surface temperature changes due to Holocene anthropogenic land cover change.
29 *Geophys. Res. Lett.*, **41**, 623–631, doi:10.1002/2013GL058085.
- 30 Heald, C. L., and D. V. Spracklen, 2015: Land Use Change Impacts on Air Quality and Climate. *Chem. Rev.*,
31 **115**, 4476–4496, doi:10.1021/cr500446g.
- 32 Heald, C. L., and J. A. Geddes, 2016: The impact of historical land use change from 1850 to 2000 on
33 secondary particulate matter and ozone. *Atmos. Chem. Phys.*, **16**, 14997–15010, doi:10.5194/acp-16-
34 14997-2016.
- 35 Heck, V., D. Gerten, W. Lucht, and A. Popp, 2018: Biomass-based negative emissions difficult to reconcile
36 with planetary boundaries. *Nat. Clim. Chang.*, doi:10.1038/s41558-017-0064-y.
- 37 Hedenus, F., S. Wirsenius, and D. J. a. Johansson, 2014: The importance of reduced meat and dairy
38 consumption for meeting stringent climate change targets. *Clim. Change*, **124**, 79–91,
39 doi:10.1007/s10584-014-1104-5.
- 40 Hegerl, G. C., and Coauthors, 2015: Challenges in quantifying changes in the global water cycle. *Bull. Am.*
41 *Meteorol. Soc.*, **96**, 1097–1115, doi:10.1175/BAMS-D-13-00212.1.
42 <http://journals.ametsoc.org/doi/10.1175/BAMS-D-13-00212.1> (Accessed November 1, 2018).
- 43 —, S. Brönnimann, A. Schurer, and T. Cowan, 2018: The early 20th century warming: Anomalies, causes,
44 and consequences. *Wiley Interdiscip. Rev. Clim. Chang.*, **9**, e522, doi:10.1002/wcc.522.
45 <http://doi.wiley.com/10.1002/wcc.522> (Accessed April 1, 2019).
- 46 Heisler-White, J. L., J. M. Blair, E. F. Kelly, K. Harmony, and A. K. Knapp, 2009: Contingent productivity
47 responses to more extreme rainfall regimes across a grassland biome. *Glob. Chang. Biol.*, **15**, 2894–
48 2904, doi:10.1111/j.1365-2486.2009.01961.x. <http://doi.wiley.com/10.1111/j.1365-2486.2009.01961.x>
49 (Accessed November 2, 2018).
- 50 Heisler, J. L., and A. K. Knapp, 2008: Temporal coherence of aboveground net primary productivity in
51 mesic grasslands. *Ecography (Cop.)*, **31**, 408–416, doi:10.1111/j.0906-7590.2008.05351.x.
52 <http://doi.wiley.com/10.1111/j.0906-7590.2008.05351.x> (Accessed November 2, 2018).
- 53 Henderson-Sellers, A., and V. Gornitz, 1984: Possible climatic impacts of land cover transformations, with
54 particular emphasis on tropical deforestation. *Clim. Change*, **6**, 231–257.
- 55 Henderson, B. B., P. J. Gerber, T. E. Hilinski, A. Falcucci, D. S. Ojima, M. Salvatore, and R. T. Conant,
56 2015a: Greenhouse gas mitigation potential of the world’s grazing lands: Modeling soil carbon and

- 1 nitrogen fluxes of mitigation practices. *Agric. Ecosyst. Environ.*, **207**, 91–100,
2 doi:10.1016/j.agee.2015.03.029.
3 <http://www.sciencedirect.com/science/article/pii/S0167880915001139>.
- 4 —, —, —, —, —, —, and —, 2015b: Greenhouse gas mitigation potential of the world's
5 grazing lands: Modeling soil carbon and nitrogen fluxes of mitigation practices. *Agric. Ecosyst.*
6 *Environ.*, **207**, 91–100, doi:10.1016/j.agee.2015.03.029.
7 <http://www.sciencedirect.com/science/article/pii/S0167880915001139>.
- 8 Hergoualc'h, K. A., and L. V. Verchot, 2012: Changes in soil CH₄ fluxes from the conversion of tropical
9 peat swamp forests: a meta-analysis. *J. Integr. Environ. Sci.*, **9**, 93–101,
10 doi:10.1080/1943815X.2012.679282. <http://www.scopus.com/inward/record.url?eid=2-s2.0-84863448205&partnerID=tZOtx3y1>.
- 11 Herrera-Estrada, J. E., and J. Sheffield, 2017: Uncertainties in future projections of summer droughts and
12 heat waves over the contiguous United States. *J. Clim.*, **30**, 6225–6246, doi:10.1175/JCLI-D-16-
13 0491.1. <http://journals.ametsoc.org/doi/10.1175/JCLI-D-16-0491.1> (Accessed May 21, 2018).
- 14 Herrero, M., S. Wirseni, B. Henderson, C. Rigolot, P. Thornton, P. Havlík, I. de Boer, and P. J. Gerber,
15 2015: Livestock and the Environment: What Have We Learned in the Past Decade? *Annu. Rev.*
16 *Environ. Resour.*, **40**, 177–202, doi:10.1146/annurev-environ-031113-093503.
17 <http://www.annualreviews.org/doi/10.1146/annurev-environ-031113-093503>.
- 18 —, and Coauthors, 2016a: Greenhouse gas mitigation potentials in the livestock sector. *Nat. Clim. Chang.*,
19 **IN PRESS**, 452, doi:10.1038/nclimate2925. <http://dx.doi.org/10.1038/nclimate2925>.
- 20 —, and Coauthors, 2016b: Greenhouse gas mitigation potentials in the livestock sector. *Nat. Clim. Chang.*,
21 **6**, 452–461, doi:10.1038/nclimate2925.
- 22 Heskell, M. A., and Coauthors, 2016: Convergence in the temperature response of leaf respiration across
23 biomes and plant functional types. *Proc. Natl. Acad. Sci.*, **113**, 3832–3837,
24 doi:10.1073/pnas.1520282113. <http://www.pnas.org/lookup/doi/10.1073/pnas.1520282113>.
- 25 Hiç, C., P. Pradhan, D. Rybski, and J. P. Kropp, 2016: Food Surplus and Its Climate Burdens. *Environ. Sci.*
26 *Technol.*, doi:10.1021/acs.est.5b05088.
- 27 Hicks Pries, C. E., C. Castanha, R. C. Porras, and M. S. Torn, 2017: The whole-soil carbon flux in response
28 to warming. *Science (80-.)*, **355**, 1420 LP-1423.
- 29 Hinkel, K. M., and F. E. Nelson, 2007: Anthropogenic heat island at Barrow, Alaska, during winter: 2001-
30 2005. *J. Geophys. Res. Atmos.*, **112**, 2001–2005, doi:10.1029/2006JD007837.
- 31 Hirsch, A. L., M. Wilhelm, E. L. Davin, W. Thiery, and S. I. Seneviratne, 2017: Can climate-effective land
32 management reduce regional warming? *J. Geophys. Res.*, **122**, 2269–2288,
33 doi:10.1002/2016JD026125. <https://doi.org/10.1002/2016JD026125>.
- 34 Hirsch, A. L., and Coauthors, 2018: Biogeophysical Impacts of Land-Use Change on Climate Extremes in
35 Low-Emission Scenarios: Results From HAPPI-Land. *Earth's Futur.*, **6**, 396–409,
36 doi:10.1002/2017EF000744.
- 37 Hirschi, M., B. Mueller, W. Dorigo, and S. I. Seneviratne, 2014: Using remotely sensed soil moisture for
38 land-atmosphere coupling diagnostics: The role of surface vs. root-zone soil moisture variability.
39 *Remote Sens. Environ.*, **154**, 246–252, doi:10.1016/j.rse.2014.08.030.
- 40 Hodzic, A., P. S. Kasibhatla, D. S. Jo, C. D. Cappa, J. L. Jimenez, S. Madronich, and R. J. Park, 2016:
41 Rethinking the global secondary organic aerosol (SOA) budget: stronger production, faster removal,
42 shorter lifetime. *Atmos. Chem. Phys.*, **16**, 7917–7941, doi:10.5194/acp-16-7917-2016.
- 43 Hoegh-Guldberg, O., and Coauthors, 2018: Impacts of 1.5°C global warming on natural and human systems.
44 *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of 1.5 °C above*
45 *pre-industrial levels and related global greenhouse gas emission pathways, in the context of*
46 *strengthening the global response to the threat of climate change*, V. Masson-Delmotte et al., Eds.,
47 Geneva http://report.ipcc.ch/sr15/pdf/sr15_chapter3.pdf (Accessed October 29, 2018).
- 48 Hoell, A., M. Hoerling, J. Eischeid, X. W. Quan, and B. Liebmann, 2017: Reconciling theories for human
49 and natural attribution of recent East Africa drying. *J. Clim.*, **30**, 1939–1957, doi:10.1175/JCLI-D-16-
50 0558.1. <http://journals.ametsoc.org/doi/10.1175/JCLI-D-16-0558.1> (Accessed October 28, 2018).
- 51 Hoerling, M., J. Eischeid, and J. Perlwitz, 2010: Regional precipitation trends: Distinguishing natural
52 variability from anthropogenic forcing. *J. Clim.*, **23**, 2131–2145, doi:10.1175/2009JCLI3420.1.
53 <http://journals.ametsoc.org/doi/abs/10.1175/2009JCLI3420.1> (Accessed May 8, 2018).
- 54 Hoesly, R. M., and Coauthors, 2018: Historical (1750–2014) anthropogenic emissions of reactive gases and
55 aerosols from the Community Emissions Data System (CEDS). *Geosci. Model Dev.*, **11**, 369–408,
56

- 1 doi:10.5194/gmd-11-369-2018. <https://www.geosci-model-dev.net/11/369/2018/>.
- 2 Hoffman, F. M., and Coauthors, 2014: Causes and implications of persistent atmospheric carbon dioxide
3 biases in Earth System Models. *J. Geophys. Res. Biogeosciences*, **119**, 141–162,
4 doi:10.1002/2013JG002381.
- 5 Hoffmann, T., S. Mudd, ... K. V. O.-E. S., and undefined 2013, Humans and the missing C-sink: erosion
6 and burial of soil carbon through time. *earth-surf-dynam.net*,.
- 7 Hollinger, D. Y., and Coauthors, 2010: Albedo estimates for land surface models and support for a new
8 paradigm based on foliage nitrogen concentration. *Glob. Chang. Biol.*, **16**, 696–710,
9 doi:10.1111/j.1365-2486.2009.02028.x.
- 10 Hooijer, A., S. Page, J. G. Canadell, M. Silvius, J. Kwadijk, H. Wösten, and J. Jauhiainen, 2010: Current and
11 future CO₂ emissions from drained peatlands in Southeast Asia. *Biogeosciences*, **7**, 1505–1514,
12 doi:10.5194/bg-7-1505-2010. <http://www.biogeosciences.net/7/1505/2010/>.
- 13 Hopkins, F. M., T. R. Filley, G. Gleixner, M. Lange, S. M. Top, and S. E. Trumbore, 2014: Increased
14 belowground carbon inputs and warming promote loss of soil organic carbon through complementary
15 microbial responses. *Soil Biol. Biochem.*, **76**, 57–69, doi:10.1016/j.soilbio.2014.04.028.
16 <http://www.sciencedirect.com/science/article/pii/S0038071714001485>.
- 17 Houghton, R. A., and A. A. Nassikas, 2017: Global and regional fluxes of carbon from land use and land
18 cover change 1850–2015. *Global Biogeochem. Cycles*, **31**, 456–472, doi:10.1002/2016GB005546.
- 19 Houghton, R. A., and A. A. Nassikas, 2018: Negative emissions from stopping deforestation and forest
20 degradation, globally. *Glob. Chang. Biol.*, **24**, 350–359, doi:10.1111/gcb.13876.
- 21 Houghton, R. A., J. I. House, J. Pongratz, G. R. Van Der Werf, R. S. Defries, M. C. Hansen, C. Le Quéré,
22 and N. Ramankutty, 2012a: Carbon emissions from land use and land-cover change. *Biogeosciences*, **9**,
23 5125–5142, doi:10.5194/bg-9-5125-2012.
- 24 —, G. R. van der Werf, R. S. DeFries, M. C. Hansen, J. I. House, C. Le Quéré, J. Pongratz, and N.
25 Ramankutty, 2012b: Chapter G2 Carbon emissions from land use and land-cover change.
26 *Biogeosciences Discuss.*, **9**, 835–878, doi:10.5194/bgd-9-835-2012.
- 27 —, B. Byers, and A. A. Nassikas, 2015: A role for tropical forests in stabilizing atmospheric CO₂. *Nat.*
28 *Clim. Chang.*, doi:10.1038/nclimate2869.
- 29 Houghton, R. A., A. Baccini, and W. S. Walker, 2018: Where is the residual terrestrial carbon sink? *Glob.*
30 *Chang. Biol.*, **24**, 3277–3279, doi:10.1111/gcb.14313. <http://doi.wiley.com/10.1111/gcb.14313>.
- 31 Houlton, B. Z., S. L. Morford, and R. A. Dahlgren, 2018: Convergent evidence for widespread rock nitrogen
32 sources in Earth's surface environment. *Science (80-.)*, **360**, 58–62, doi:10.1126/science.aan4399.
- 33 Houweling, S., and Coauthors, 2014: A multi-year methane inversion using SCIAMACHY, accounting for
34 systematic errors using TCCON measurements. *Atmos. Chem. Phys.*, **14**, 3991–4012, doi:10.5194/acp-
35 14-3991-2014.
- 36 Howard, J., A. Sutton-Grier, D. Herr, J. Kleypas, E. Landis, E. Mcleod, E. Pidgeon, and S. Simpson, 2017:
37 Clarifying the role of coastal and marine systems in climate mitigation. *Front. Ecol. Environ.*, **15**, 42–
38 50, doi:10.1002/fee.1451. <https://doi.org/10.1002/fee.1451>.
- 39 Howden, S. M., J.-F. Soussana, F. N. Tubiello, N. Chhetri, M. Dunlop, and H. Meinke, 2007: Adapting
40 agriculture to climate change. *Proc. Natl. Acad. Sci.*, **104**, 19691–19696,
41 doi:10.1073/pnas.0701890104. <http://www.pnas.org/cgi/doi/10.1073/pnas.0701890104>.
- 42 Hristov, A. N., and Coauthors, 2013: *FAO Animal Production and Health Paper No 177 BT - Mitigation of*
43 *Greenhouse Gas Emissions in Livestock Production: A Review of Technical Options for Non-CO₂*
44 *Emissions*. Rome, 226pp pp.
- 45 Hu, W. W., and Coauthors, 2015: Characterization of a real-time tracer for isoprene epoxydiols-derived
46 secondary organic aerosol (IEPOX-SOA) from aerosol mass spectrometer measurements. *Atmos.*
47 *Chem. Phys.*, **15**, 11807–11833, doi:10.5194/acp-15-11807-2015.
- 48 Hua, W., H. Chen, S. Sun, and L. Zhou, 2015: Assessing climatic impacts of future land use and land cover
49 change projected with the CanESM2 model. *Int. J. Climatol.*, doi:10.1002/joc.4240.
- 50 Hua, W. J., and H. S. Chen, 2013: Recognition of climatic effects of land use/land cover change under global
51 warming. *Chinese Sci. Bull.*, doi:10.1007/s11434-013-5902-3.
- 52 Huang, J., and Coauthors, 2017: Dryland climate change: Recent progress and challenges. *Rev. Geophys.*,
53 **55**, 719–778, doi:10.1002/2016RG000550. <http://doi.wiley.com/10.1002/2016RG000550> (Accessed
54 May 24, 2018).
- 55 Huang, W., and S. J. Hall, 2017: Elevated moisture stimulates carbon loss from mineral soils by releasing
56 protected organic matter. *Nat. Commun.*, **8**, 1774, doi:10.1038/s41467-017-01998-z.

- 1 <https://doi.org/10.1038/s41467-017-01998-z>.
- 2 Huang, Y., S. Wu, and J. O. Kaplan, 2014: Sensitivity of global wildfire occurrences to various factors in
3 the context of global change. *Atmos. Environ.*, **121**, 86–92, doi:10.1016/j.atmosenv.2015.06.002.
4 <http://www.sciencedirect.com/science/article/pii/S1352231015301448>.
- 5 Huber, D., D. Mechem, and N. Brunzell, 2014: The Effects of Great Plains Irrigation on the Surface Energy
6 Balance, Regional Circulation, and Precipitation. *Climate*, **2**, 103–128, doi:10.3390/cli2020103.
- 7 Hudiburg, T. W., B. E. Law, C. Wirth, and S. Luyssaert, 2011: Regional carbon dioxide implications of
8 forest bioenergy production. *Nat. Clim. Chang.*, **1**, 419–423,
9 doi:<http://www.nature.com/nclimate/journal/v1/n8/abs/nclimate1264.html#supplementary-information>.
10 <http://dx.doi.org/10.1038/nclimate1264>.
- 11 Huffman, J. A., and Coauthors, 2012: Size distributions and temporal variations of biological aerosol
12 particles in the Amazon rainforest characterized by microscopy and real-time UV-APS fluorescence
13 techniques during AMAZE-08. *Atmos. Chem. Phys.*, **12**, 11997–12019, doi:10.5194/acp-12-11997-
14 2012.
- 15 Hugelius, G., C. Tarnocai, G. Broll, J. G. Canadell, P. Kuhry, and D. K. Swanson, 2013: The northern
16 circumpolar soil carbon database: Spatially distributed datasets of soil coverage and soil carbon storage
17 in the northern permafrost regions. *Earth Syst. Sci. Data*, **5**, 3–13, doi:10.5194/essd-5-3-2013.
18 <https://www.earth-syst-sci-data.net/5/3/2013/>.
- 19 —, and Coauthors, 2014: Estimated stocks of circumpolar permafrost carbon with quantified uncertainty
20 ranges and identified data gaps. *Biogeosciences*, **11**, 6573–6593, doi:10.5194/bg-11-6573-2014.
- 21 Huijnen, V., and Coauthors, 2016: Fire carbon emissions over maritime southeast Asia in 2015 largest since
22 1997. *Sci. Rep.*, **6**, 26886, doi:10.1038/srep26886. <http://www.nature.com/articles/srep26886>.
- 23 Hulme, M., 2001: Climatic perspectives on Sahelian desiccation: 1973–1998. *Glob. Environ. Chang.*, **11**, 19–
24 29, doi:10.1016/S0959-3780(00)00042-X.
- 25 —, E. M. Barrow, N. W. Arnell, P. A. Harrison, T. C. Johns, and T. E. Downing, 1999: Relative impacts
26 of human-induced climate change and natural climate variability. *Nature*, **397**, 688–691,
27 doi:10.1038/17789. <http://www.nature.com/articles/17789> (Accessed May 7, 2018).
- 28 Humpenöder, F., and Coauthors, 2018: Large-scale bioenergy production: how to resolve sustainability
29 trade-offs? *Environ. Res. Lett.*, **13**, 24011, doi:10.1088/1748-9326/aa9e3b.
- 30 Humphrey, V., J. Zscheischler, P. Ciais, L. Gudmundsson, S. Sitch, and S. I. Seneviratne, 2018: Sensitivity
31 of atmospheric CO₂ growth rate to observed changes in terrestrial water storage. *Nature*, **560**, 628–631,
32 doi:10.1038/s41586-018-0424-4. <http://www.nature.com/articles/s41586-018-0424-4>.
- 33 Hungate, B. A., and Coauthors, 2013: Cumulative response of ecosystem carbon and nitrogen stocks to
34 chronic CO₂ exposure in a subtropical oak woodland. *New Phytol.*, **200**, 753–766,
35 doi:10.1111/nph.12333. <https://doi.org/10.1111/nph.12333>.
- 36 Huntingford, C., and Coauthors, 2017: Implications of improved representations of plant respiration in a
37 changing climate. *Nat. Commun.*, **8**, 1602, doi:10.1038/s41467-017-01774-z.
38 <http://www.nature.com/articles/s41467-017-01774-z>.
- 39 Huppmann, D., J. Rogelj, E. Kriegler, V. Krey, and K. Riahi, 2018: A new scenario resource for integrated
40 1.5 °C research. *Nat. Clim. Chang.*, **8**, 1027–1030, doi:10.1038/s41558-018-0317-4.
- 41 Hurtt, G., L. Chini, R. Sahajpal, and S. Frolking, 2017: Harmonization of global land-use change and
42 management for the period 850–2100. *Geosci. Model Dev.*, <http://luh.umd.edu/>.
- 43 Hurtt, G. C., and Coauthors, 2011: Harmonization of land-use scenarios for the period 1500–2100: 600 years
44 of global gridded annual land-use transitions, wood harvest, and resulting secondary lands. *Clim.*
45 *Change*, **109**, 117–161, doi:10.1007/s10584-011-0153-2.
- 46 Hussain, S., S. Peng, S. Fahad, A. Khaliq, J. Huang, K. Cui, and L. Nie, 2015: Rice management
47 interventions to mitigate greenhouse gas emissions: a review. *Environ. Sci. Pollut. Res.*, **22**, 3342–
48 3360, doi:10.1007/s11356-014-3760-4. <https://doi.org/10.1007/s11356-014-3760-4>.
- 49 IFPRI, 2009: Climate change: Impact on agriculture and costs of adaptation. *Int. Food Policy*, 307–324,
50 doi:10.2499/0896295354.
- 51 Im, E. S., and E. A. B. Eltahir, 2014: Enhancement of rainfall and runoff upstream from irrigation location in
52 a climate model of West Africa. *Water Resour. Res.*, **50**, 8651–8674, doi:10.1002/2014WR015592.
- 53 —, M. P. Marcella, and E. A. B. Eltahir, 2014: Impact of potential large-scale irrigation on the West
54 African monsoon and its dependence on location of irrigated area. *J. Clim.*, **27**, 994–1009,
55 doi:10.1175/JCLI-D-13-00290.1.
- 56 Intergovernmental Panel on Climate Change, 2014: Human Settlements, Infrastructure, and Spatial Planning.

- 1 *Climate Change 2014 Mitigation of Climate Change*, O. Edenhofer et al., Eds., Cambridge University
2 Press, Cambridge, United Kingdom and New York, NY, USA, 923–1000
3 <http://ebooks.cambridge.org/ref/id/CBO9781107415416A026>.
- 4 Iordan, C. M., X. Hu, A. Arvesen, P. Kauppi, and F. Cherubini, 2018: Contribution of forest wood products
5 to negative emissions: Historical comparative analysis from 1960 to 2015 in Norway, Sweden and
6 Finland. *Carbon Balance Manag.*, doi:10.1186/s13021-018-0101-9.
- 7 IPCC, 2003: *Good Practice Guidance for Land Use, and-Use Change and Forestry*. J. Penman et al., Eds.
8 Institute for Global Environmental Strategies (IGES) for the IPCC, Hayama, Kanagawa, 590 pp.
- 9 ———, 2005: *Special Report on Carbon Dioxide Capture and Storage. Prepared by Working Group III of the*
10 *Intergovernmental Panel on Climate Change*. B. Metz, O. Davidson, H.C. De Coninck, M. Loos, and
11 L.A. Meyer, Eds. Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.,
12 ———, 2006: 2006 IPCC Guidelines for National Greenhouse Gas Inventories. Institute for Global
13 Environmental Strategies (IGES) for the Intergovernmental Panel on Climate Change.
- 14 ———, 2010a: Revisiting the Use of Managed Land as a Proxy for Estimating National Anthropogenic
15 Emissions and Removals. S. et al. Eggleston, Ed., IGES.
- 16 ———, 2010b: Revisiting the use of managed land as a proxy for estimating national anthropogenic emissions
17 and removals. *Meeting Report, 5-7 May, 2009, INPE, Sao Jose dos Campos, Brazil.*, H.S. Eggleston, N.
18 Srivastava, K. Tanabe, and J. Baasansuren, Eds., Institute for Global Environmental Strategies (IGES).
- 19 ———, 2013a: Summary for Policymakers. *CLIMATE CHANGE 2013 - The Physical Science Basis,*
20 *Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel on*
21 *Climate Change*, C.B. Field et al., Eds., Cambridge University Press, Cambridge, United Kingdom, and
22 New York, NY, USA, p. 78 http://report.ipcc.ch/sr15/pdf/sr15_spm_final.pdf (Accessed October 29,
23 2018).
- 24 ———, 2013b: AR5 - Summary for Policymakers. *Clim. Chang. 2013 - Phys. Sci. Basis, Contrib. Work. Gr. I*
25 *to Fifth Assess. Rep. Intergov. Panel Clim. Chang.*, 78, doi:10.1017/CBO9781107415324.Summary.
- 26 ———, 2013c: *Climate Change 2013: The Physical Science Basis*. T.F. Stocker et al., Eds. Cambridge
27 University Press, Cambridge, United Kingdom and New York, NY, USA.,
- 28 ———, 2014: Climate Change 2013: The Physical Science Basis. *Climate Change 2013: The Physical Science*
29 *Basis. Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel*
30 *on Climate Change*, T.F. Stocker et al., Eds., Cambridge University Press, Cambridge, United
31 Kingdom and New York, NY, USA, United Kingdom and New York, NY, USA, 1523–1535.
- 32 ———, 2018: Summary for Policymakers. *Global Warming of 1.5°C. An IPCC Special Report on the impacts*
33 *of global warming of 1.5°C above pre-industrial levels and related global greenhouse gas emission*
34 *pathways, in the context of strengthening the global response to the threat of climate change*, V.
35 Masson-Delmotte et al., Eds., World Meteorological Organization, Geneva, Switzerland, Geneva, p.
36 26.
- 37 Ise, T., A. L. Dunn, S. C. Wofsy, and P. R. Moorcroft, 2008: High sensitivity of peat decomposition to
38 climate change through water-table feedback. *Nat. Geosci.*, doi:10.1038/ngeo331.
- 39 Ito, A., K. Nishina, and H. M. Noda, 2016: Impacts of future climate change on the carbon budget of
40 northern high-latitude terrestrial ecosystems: An analysis using ISI-MIP data. *Polar Sci.*, **10**, 346–355,
41 doi:10.1016/j.polar.2015.11.002.
- 42 Jacinthe, P. A., and R. Lal, 2001: A mass balance approach to assess carbon dioxide evolution during
43 erosional events. *L. Degrad. Dev.*, **12**, 329–339, doi:doi:10.1002/ldr.454.
44 <https://doi.org/10.1002/ldr.454>.
- 45 Jackson, R. B., and Coauthors, 2017: The Ecology of Soil Carbon: Pools, Vulnerabilities, and Biotic and
46 Abiotic Controls. *Annu. Rev. Ecol. Evol. Syst.*, **48**, annurev-ecolsys-112414-054234,
47 doi:10.1146/annurev-ecolsys-112414-054234. <http://www.annualreviews.org/doi/10.1146/annurev-ecolsys-112414-054234>.
- 48 Jacobson, M., H. Hansson, K. Noone, and R. Charlson, 2000: Organic atmospheric aerosols: Review and
49 state of the science. *Rev. Geophys.*, **38**, 267–294.
- 50 Janssens-Maenhout, G., and Coauthors, 2017a: EDGAR v4.3.2 Global Atlas of the three major Greenhouse
51 Gas Emissions for the period 1970--2012. *Earth Syst. Sci. Data Discuss.*, **2017**, 1–55,
52 doi:10.5194/essd-2017-79.
- 53 Janssens-Maenhout, G., M. Crippa, D. Guizzardi, M. Muntean, E. Schaaf, J. G. J. Olivier, J. A. H. W. Peters,
54 and K. M. Schure, 2017b: *Fossil CO₂ & GHG emissions of all world countries*. 239 pp.
- 55 Janssens, I. A., and Coauthors, 2010: Reduction of forest soil respiration in response to nitrogen deposition.

- 1 *Nat. Geosci.*, **3**, 315–322, doi:10.1038/ngeo844.
- 2 Jardine, K., and Coauthors, 2011: Within-canopy sesquiterpene ozonolysis in Amazonia. *J. Geophys. Res.*
3 *Atmos.*, **116**, 1–10, doi:10.1029/2011JD016243.
- 4 Jardine, K., and Coauthors, 2015: Green Leaf Volatile Emissions during High Temperature and Drought
5 Stress in a Central Amazon Rainforest. *Plants*, **4**, 678–690, doi:10.3390/plants4030678.
- 6 Jennerjahn, T. C., E. Gilman, K. W. Krauss, L. D. Lacerda, I. Nordhaus, and E. Wolanski, 2017: Mangrove
7 Ecosystems under Climate Change. *Mangrove Ecosystems: A Global Biogeographic Perspective*, V.H.
8 Rivera-Monroy, S. Y. Lee, E. Kristensen, and R.R. Twilley, Eds., Springer International Publishing,
9 Cham, 211–244 http://link.springer.com/10.1007/978-3-319-62206-4_7.
- 10 Jeong, J. H., J. S. Kug, H. W. Linderholm, D. Chen, B. M. Kim, and S. Y. Jun, 2014a: Intensified Arctic
11 warming under greenhouse warming by vegetation-atmosphere-sea ice interaction. *Environ. Res. Lett.*,
12 **9**, doi:10.1088/1748-9326/9/9/094007.
- 13 Jeong, S. J., C. H. Ho, S. Piao, J. Kim, P. Ciais, Y. B. Lee, J. G. Jhun, and S. K. Park, 2014b: Effects of
14 double cropping on summer climate of the North China Plain and neighbouring regions. *Nat. Clim.*
15 *Chang.*, **4**, 615–619, doi:10.1038/nclimate2266.
- 16 Jiang, L., 2014: Internal consistency of demographic assumptions in the shared socioeconomic pathways.
17 *Popul. Environ.*, **35**, 261–285, doi:10.1007/s11111-014-0206-3. [https://doi.org/10.1007/s11111-014-](https://doi.org/10.1007/s11111-014-0206-3)
18 [0206-3](https://doi.org/10.1007/s11111-014-0206-3).
- 19 Johnstone, J. F., and Coauthors, 2016: Changing disturbance regimes, ecological memory, and forest
20 resilience. *Front. Ecol. Environ.*, **14**, 369–378, doi:10.1002/fee.1311.
21 <http://doi.wiley.com/10.1002/fee.1311> (Accessed October 30, 2018).
- 22 Jolly, W. M., M. A. Cochrane, P. H. Freeborn, Z. A. Holden, T. J. Brown, G. J. Williamson, and D. M. J. S.
23 Bowman, 2015: Climate-induced variations in global wildfire danger from 1979 to 2013. *Nat.*
24 *Commun.*, **6**, 7537, doi:10.1038/ncomms8537. <http://www.nature.com/doifinder/10.1038/ncomms8537>.
- 25 Jonard, M., and Coauthors, 2015: Tree mineral nutrition is deteriorating in Europe. *Glob. Chang. Biol.*, **21**,
26 418–430, doi:10.1111/gcb.12657.
- 27 Jones, A. D., and Coauthors, 2013a: Greenhouse gas policy influences climate via direct effects of land-use
28 change. *J. Clim.*, **26**, 3657–3670, doi:10.1175/JCLI-D-12-00377.1.
- 29 ———, W. D. Collins, and M. S. Torn, 2013b: On the additivity of radiative forcing between land use change
30 and greenhouse gases. *Geophys. Res. Lett.*, doi:10.1002/grl.50754.
- 31 Jones, C., J. Lowe, S. Liddicoat, and R. Betts, 2009: Committed terrestrial ecosystem changes due to climate
32 change. *Nat. Geosci.*, **2**, 484–487, doi:10.1038/ngeo555. <http://www.nature.com/articles/ngeo555>
33 (Accessed March 13, 2019).
- 34 ———, S. Liddicoat, and J. Lowe, 2010: Role of terrestrial ecosystems in determining CO₂ stabilization and
35 recovery behaviour. *Tellus, Ser. B Chem. Phys. Meteorol.*, **62**, 682–699, doi:10.1111/j.1600-
36 0889.2010.00490.x.
- 37 Jones, C. D., and Coauthors, 2016: Simulating the Earth system response to negative emissions. *Environ.*
38 *Res. Lett.*, **11**, 95012, doi:10.1088/1748-9326/11/9/095012. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=9/a=095012?key=crossref.6b5747055a178d1c59ffa940adb33091)
39 [9326/11/i=9/a=095012?key=crossref.6b5747055a178d1c59ffa940adb33091](http://stacks.iop.org/1748-9326/11/i=9/a=095012?key=crossref.6b5747055a178d1c59ffa940adb33091).
- 40 Jones, P. D., D. H. Lister, T. J. Osborn, C. Harpham, M. Salmon, and C. P. Morice, 2012: Hemispheric and
41 large-scale land-surface air temperature variations: An extensive revision and an update to 2010. *J.*
42 *Geophys. Res. Atmos.*, **117**, doi:10.1029/2011JD017139. <https://doi.org/10.1029/2011JD017139>.
- 43 de Jong, R., J. Verbesselt, M. E. Schaepman, and S. de Bruin, 2012: Trend changes in global greening and
44 browning: Contribution of short-term trends to longer-term change. *Glob. Chang. Biol.*, **18**, 642–655,
45 doi:10.1111/j.1365-2486.2011.02578.x.
- 46 Joosten, H., 2015: Current soil carbon loss and land degradation globally-where are the hotspots and why
47 there? *Soil carbon: science, management and policy for multiple benefits*, S.A. Banwart, E.
48 Noellemeyer, and E. Milne, Eds., Wallingford, UK, UK, 224–234.
- 49 Jørgensen, S. E., L. Xu, J. Marques, and F. Salas, 2016: Application of indicators for the assessment of
50 ecosystem health. *Handbook of ecological indicators for assessment of ecosystem health*, S.E.
51 Jørgensen, L. Xu, and R. Costanza, Eds., CRC Press, Taylor & Francis Group, 22–89.
- 52 Joshi, M. M., F. H. Lambert, and M. J. Webb, 2013: An explanation for the difference between twentieth and
53 twenty-first century land-sea warming ratio in climate models. *Clim. Dyn.*, **41**, 1853–1869,
54 doi:10.1007/s00382-013-1664-5.
- 55 Journet, E., Y. Balkanski, and S. P. Harrison, 2014: A new data set of soil mineralogy for dust-cycle
56 modeling. *Atmos. Chem. Phys.*, **14**, 3801–3816, doi:10.5194/acp-14-3801-2014.

- 1 Juang, J. Y., G. Katul, M. Siqueira, P. Stoy, and K. Novick, 2007: Separating the effects of albedo from eco-
2 physiological changes on surface temperature along a successional chronosequence in the southeastern
3 United States. *Geophys. Res. Lett.*, **34**, 1–5, doi:10.1029/2007GL031296.
- 4 Jylhä, K., H. Tuomenvirta, K. Ruostenoja, H. Niemi-Hugaerts, K. Keisu, and J. A. Karhu, 2010: Observed
5 and Projected Future Shifts of Climatic Zones in Europe and Their Use to Visualize Climate Change
6 Information. *Weather. Clim. Soc.*, **2**, 148–167, doi:10.1175/2010wcas1010.1.
7 <http://journals.ametsoc.org/doi/abs/10.1175/2010WCAS1010.1> (Accessed April 15, 2019).
- 8 Kageyama, M., S. Charbit, C. Ritz, M. Khodri, and G. Ramstein, 2004: Quantifying ice-sheet feedbacks
9 during the last glacial inception. *Geophys. Res. Lett.*, **31**, 1–4, doi:10.1029/2004GL021339.
- 10 Kallenbach, C. M., S. D. Frey, and A. S. Grandy, 2016: Direct evidence for microbial-derived soil organic
11 matter formation and its ecophysiological controls. *Nat. Commun.*, **7**, 13630,
12 doi:10.1038/ncomms13630. <http://dx.doi.org/10.1038/ncomms13630>.
- 13 Karhu, K., and Coauthors, 2014: Temperature sensitivity of soil respiration rates enhanced by microbial
14 community response. *Nature*, **513**, 81–84, doi:10.1038/nature13604.
- 15 Kattge, J., and Coauthors, 2007: Temperature acclimation in a biochemical model of photosynthesis: A
16 reanalysis of data from 36 species. *Plant, Cell Environ.*, **30**, 1176–1190, doi:10.1111/j.1365-
17 3040.2007.01690.x.
- 18 De Kauwe, M. G., T. F. Keenan, B. E. Medlyn, I. C. Prentice, and C. Terrer, 2016: Satellite based estimates
19 underestimate the effect of CO₂ fertilization on net primary productivity. *Nat. Clim. Chang.*, **6**, 892–
20 893, doi:10.1038/nclimate3105.
- 21 Kaye, J. P., and M. Quemada, 2017: Using cover crops to mitigate and adapt to climate change . A review.
22 doi:10.1007/s13593-016-0410-x.
- 23 Keenan, T. F., and W. J. Riley, 2018: Greening of the land surface in the world’s cold regions consistent
24 with recent warming. *Nat. Clim. Chang.*, **8**, 825–828, doi:10.1038/s41558-018-0258-y.
25 <http://www.nature.com/articles/s41558-018-0258-y> (Accessed October 16, 2018).
- 26 Keenan, T. F., D. Y. Hollinger, G. Bohrer, D. Dragoni, J. W. Munger, H. P. Schmid, and A. D. Richardson,
27 2013: Increase in forest water-use efficiency as atmospheric carbon dioxide concentrations rise. *Nature*,
28 **499**, 324–327, doi:10.1038/nature12291.
- 29 Keenan, T. F., I. C. Prentice, J. G. Canadell, C. A. Williams, H. Wang, M. Raupach, and G. J. Collatz, 2016:
30 Recent pause in the growth rate of atmospheric CO₂ due to enhanced terrestrial carbon uptake. *Nat.*
31 *Commun.*, **7**, doi:10.1038/ncomms13428.
- 32 Keenan, T. F., I. C. Prentice, J. G. Canadell, C. A. Williams, H. Wang, M. Raupach, and G. J. Collatz, 2017:
33 Corrigendum: Recent pause in the growth rate of atmospheric CO₂ due to enhanced terrestrial carbon
34 uptake. *Nat. Commun.*, **8**, 16137, doi:10.1038/ncomms16137.
- 35 Keiluweit, M., J. J. Bougoure, P. S. Nico, J. Pett-Ridge, P. K. Weber, and M. Kleber, 2015: Mineral
36 protection of soil carbon counteracted by root exudates. *Nat. Clim. Chang.*, **5**, 588–595,
37 doi:10.1038/nclimate2580. <http://dx.doi.org/10.1038/nclimate2580>.
- 38 ———, P. S. Nico, M. Kleber, and S. Fendorf, 2016: Are oxygen limitations under recognized regulators of
39 organic carbon turnover in upland soils? *Biogeochemistry*, **127**, 157–171, doi:10.1007/s10533-015-
40 0180-6. <https://doi.org/10.1007/s10533-015-0180-6>.
- 41 Keith, H., D. Lindenmayer, B. Mackey, D. Blair, L. Carter, L. McBurney, S. Okada, and T. Konishi-Nagano,
42 2014: Managing temperate forests for carbon storage: impacts of logging versus forest protection on
43 carbon stocks. *Ecosphere*, **5**, art75, doi:10.1890/ES14-00051.1.
- 44 Kelleway, J. J., K. Cavanaugh, K. Rogers, I. C. Feller, E. Ens, C. Doughty, and N. Saintilan, 2017: Review
45 of the ecosystem service implications of mangrove encroachment into salt marshes. *Glob. Chang. Biol.*,
46 **23**, 3967–3983, doi:10.1111/gcb.13727. <https://doi.org/10.1111/gcb.13727>.
- 47 Kelley, C. P., S. Mohtadi, M. A. Cane, R. Seager, and Y. Kushnir, 2015: Climate change in the Fertile
48 Crescent and implications of the recent Syrian drought. *Proc. Natl. Acad. Sci.*, **112**, 3241–3246,
49 doi:10.1073/pnas.1421533112. <http://www.ncbi.nlm.nih.gov/pubmed/25733898> (Accessed September
50 29, 2018).
- 51 Kelly, R., M. L. Chipman, P. E. Higuera, I. Stefanova, L. B. Brubaker, and F. S. Hu, 2013: Recent burning
52 of boreal forests exceeds fire regime limits of the past 10,000 years. *Proc. Natl. Acad. Sci.*, **110**, 13055–
53 13060, doi:10.1073/pnas.1305069110.
- 54 Kemena, T. P., K. Matthes, T. Martin, S. Wahl, and A. Oschlies, 2017: Atmospheric feedbacks in North
55 Africa from an irrigated, afforested Sahara. *Clim. Dyn.*, **0**, 1–21, doi:10.1007/s00382-017-3890-8.
- 56 Kendon, E. J., and Coauthors, 2017: Do Convection-Permitting Regional Climate Models Improve

- 1 Projections of Future Precipitation Change? *Bull. Am. Meteorol. Soc.*, **98**, 79–93, doi:10.1175/BAMS-
2 D-15-0004.1.
- 3 Kendra Gotangco Castillo, C., and K. R. Gurney, 2013: A sensitivity analysis of surface biophysical, carbon,
4 and climate impacts of tropical deforestation rates in CCSM4-CNDV. *J. Clim.*, **26**, 805–821,
5 doi:10.1175/JCLI-D-11-00382.1.
- 6 van Kessel, C., R. Venterea, J. Six, M. A. Adviento-Borbe, B. Linqvist, and K. J. van Groenigen, 2013:
7 Climate, duration, and N placement determine N₂O emissions in reduced tillage systems: A meta-
8 analysis. *Glob. Chang. Biol.*, **19**, 33–44, doi:10.1111/j.1365-2486.2012.02779.x.
- 9 Kicklighter, D., and Coauthors, 2013: Climate impacts of a large-scale biofuels expansion. *Geophys. Res.*
10 *Lett.*, **40**, 1624–1630, doi:10.1002/grl.50352. <http://dx.doi.org/10.1002/grl.50352>.
- 11 Kim, D., 2014: The effect of land-use change on the net exchange rates of greenhouse gases: a meta-
12 analytical approach. doi:10.5194/bgd-11-1053-2014.
- 13 Kim, H., Y. K. Kim, S. K. Song, and H. W. Lee, 2016: Impact of future urban growth on regional climate
14 changes in the Seoul Metropolitan Area, Korea. *Sci. Total Environ.*, **571**, 355–363,
15 doi:10.1016/j.scitotenv.2016.05.046.
- 16 Kim, S., B. E. Dale, R. Heijungs, A. Azapagic, T. Darlington, and D. Kahlbaum, 2014: Indirect land use
17 change and biofuels: Mathematical analysis reveals a fundamental flaw in the regulatory approach.
18 *Biomass and Bioenergy*, **71**, 408–412, doi:<https://doi.org/10.1016/j.biombioe.2014.09.015>.
19 <http://www.sciencedirect.com/science/article/pii/S0961953414004255>.
- 20 King, A. D., 2017: Attributing Changing Rates of Temperature Record Breaking to Anthropogenic
21 Influences. *Earth's Futur.*, **5**, 1156–1168, doi:10.1002/2017EF000611.
22 <http://doi.wiley.com/10.1002/2017EF000611> (Accessed October 28, 2018).
- 23 —, M. T. Black, S. K. Min, E. M. Fischer, D. M. Mitchell, L. J. Harrington, and S. E. Perkins-Kirkpatrick,
24 2016: Emergence of heat extremes attributable to anthropogenic influences. *Geophys. Res. Lett.*, **43**,
25 3438–3443, doi:10.1002/2015GL067448.
- 26 Kleber, M., and Coauthors, 2011: Old and stable soil organic matter is not necessarily chemically
27 recalcitrant: Implications for modeling concepts and temperature sensitivity. *Glob. Chang. Biol.*, **17**,
28 1097–1107, doi:10.1111/j.1365-2486.2010.02278.x. <https://doi.org/10.1111/j.1365-2486.2010.02278.x>.
- 29 Klein, C., J. Bliefernicht, D. Heinzeller, U. Gessner, I. Klein, and H. Kunstmann, 2017: Feedback of
30 observed interannual vegetation change: a regional climate model analysis for the West African
31 monsoon. *Clim. Dyn.*, **48**, 2837–2858, doi:10.1007/s00382-016-3237-x.
- 32 de Klein, C. A. M., M. A. Shepherd, and T. J. van der Weerden, 2014: Nitrous oxide emissions from grazed
33 grasslands: Interactions between the N cycle and climate change - a New Zealand case study. *Curr.*
34 *Opin. Environ. Sustain.*, **9–10**, 131–139, doi:10.1016/j.cosust.2014.09.016.
- 35 Klemm, W., B. G. Heusinkveld, S. Lenzholzer, M. H. Jacobs, and B. Van Hove, 2015: Psychological and
36 physical impact of urban green spaces on outdoor thermal comfort during summertime in The
37 Netherlands. *Build. Environ.*, **83**, 120–128, doi:<https://doi.org/10.1016/j.buildenv.2014.05.013>.
- 38 Klimont, Z., K. Kupiainen, C. Heyes, P. Purohit, J. Cofala, P. Rafaj, J. Borken-Kleefeld, and W. Schöpp,
39 2017: Global anthropogenic emissions of particulate matter including black carbon. *Atmos. Chem.*
40 *Phys.*, **17**, 8681–8723, doi:10.5194/acp-17-8681-2017.
- 41 Kløve, B., and Coauthors, 2014: Climate change impacts on groundwater and dependent ecosystems. *J.*
42 *Hydrol.*, **518**, 250–266, doi:10.1016/j.jhydrol.2013.06.037.
43 <https://www.sciencedirect.com/science/article/pii/S0022169413004800> (Accessed November 1, 2018).
- 44 Knapp, A. K., and M. D. Smith, 2001: Variation among biomes in temporal dynamics of aboveground
45 primary production. *Science (80-.)*, **291**, 481–484, doi:10.1126/science.291.5503.481.
46 <http://www.ncbi.nlm.nih.gov/pubmed/11161201> (Accessed November 2, 2018).
- 47 Knapp, A. K., and Coauthors, 2008: Consequences of More Extreme Precipitation Regimes for Terrestrial
48 Ecosystems. *Bioscience*, **58**, 811–821, doi:10.1641/B580908.
49 [http://academic.oup.com/bioscience/article/58/9/811/250853/Consequences-of-More-Extreme-](http://academic.oup.com/bioscience/article/58/9/811/250853/Consequences-of-More-Extreme-Precipitation-Regimes)
50 [Precipitation-Regimes](http://academic.oup.com/bioscience/article/58/9/811/250853/Consequences-of-More-Extreme-Precipitation-Regimes) (Accessed November 2, 2018).
- 51 Knorr, W., I. C. Prentice, J. I. House, and E. A. Holland, 2005: Long-term sensitivity of soil carbon turnover
52 to warming. *Nature*, **433**, 298–301, doi:10.1038/nature03226.
- 53 Knorr, W., A. Arneth, and L. Jiang, 2016a: Demographic controls of future global fire risk. *Nat. Clim.*
54 *Chang.*, **6**, 781–785, doi:10.1038/nclimate2999.
- 55 —, L. Jiang, and A. Arneth, 2016b: Climate, CO₂ and human population impacts on global wildfire
56 emissions. *Biogeosciences*, **13**, 267–282, doi:10.5194/bg-13-267-2016.

- 1 Köchy, M., R. Hiederer, and A. Freibauer, 2015: Global distribution of soil organic carbon – Part 1: Masses
2 and frequency distributions of SOC stocks for the tropics, permafrost regions, wetlands, and the world.
3 *SOIL*, **1**, 351–365, doi:10.5194/soil-1-351-2015. <https://www.soil-journal.net/1/351/2015/>.
- 4 Kögel-Knabner, I., and Coauthors, 2008: Organo-mineral associations in temperate soils: Integrating
5 biology, mineralogy, and organic matter chemistry. *J. Plant Nutr. Soil Sci.*, **171**, 61–82,
6 doi:10.1002/jpln.200700048. <https://doi.org/10.1002/jpln.200700048>.
- 7 Kok, J. F., 2011: A scaling theory for the size distribution of emitted dust aerosols suggests climate models
8 underestimate the size of the global dust cycle. *Proc. Natl. Acad. Sci.*, **108**, 1016–1021,
9 doi:10.1073/pnas.1014798108.
- 10 Kok, J. F., S. Albani, N. M. Mahowald, and D. S. Ward, 2014: An improved dust emission model - Part 2:
11 Evaluation in the Community Earth System Model, with implications for the use of dust source
12 functions. *Atmos. Chem. Phys.*, **14**, 13043–13061, doi:10.5194/acp-14-13043-2014.
- 13 Kok, J. F., D. S. Ward, N. M. Mahowald, and A. T. Evan, 2018: Global and regional importance of the direct
14 dust-climate feedback. *Nat. Commun.*, **9**, 241, doi:10.1038/s41467-017-02620-y.
- 15 Kolus, H. R., and Coauthors, 2019: Land carbon models underestimate the severity and duration of drought's
16 impact on plant productivity. *Sci. Rep.*, **9**, 2758, doi:10.1038/s41598-019-39373-1.
17 <http://www.nature.com/articles/s41598-019-39373-1>.
- 18 Kooperman, G. J., M. S. Pritchard, and R. C. J. Somerville, 2014: The response of US summer rainfall to
19 quadrupled CO₂ climate change in conventional and superparameterized versions of the NCAR
20 community atmosphere model. *J. Adv. Model. Earth Syst.*, **6**, 859–882, doi:10.1002/2014MS000306.
21 <http://doi.wiley.com/10.1002/2014MS000306> (Accessed October 30, 2018).
- 22 ———, ———, T. A. O'Brien, and B. W. Timmermans, 2018: Rainfall From Resolved Rather Than
23 Parameterized Processes Better Represents the Present-Day and Climate Change Response of Moderate
24 Rates in the Community Atmosphere Model. *J. Adv. Model. Earth Syst.*, **10**, 971–988,
25 doi:10.1002/2017MS001188. <http://doi.wiley.com/10.1002/2017MS001188> (Accessed October 30,
26 2018).
- 27 Koornneef, J., and Coauthors, 2012: Global potential for biomass and carbon dioxide capture, transport and
28 storage up to 2050. *Int. J. Greenh. Gas Control*, doi:10.1016/j.ijggc.2012.07.027.
- 29 Körner, C., 2006: Plant CO₂ responses: An issue of definition, time and resource supply. *New Phytol.*, **172**,
30 393–411, doi:10.1111/j.1469-8137.2006.01886.x.
- 31 ———, 2015: Paradigm shift in plant growth control. *Curr. Opin. Plant Biol.*, **25**, 107–114,
32 doi:10.1016/j.pbi.2015.05.003.
- 33 Koskinen, M., L. Maanaviija, M. Nieminen, K. Minkkinen, and E.-S. Tuittila, 2016: High methane emissions
34 from restored Norway spruce swamps in southern Finland over one growing season. *Mires Peat*, **17**, 1–
35 13, doi:10.19189/MaP.2015.OMB.202.
- 36 Kotteck, M., J. Grieser, C. Beck, B. Rudolf, and F. Rubel, 2006: World map of the Köppen-Geiger climate
37 classification updated. *Meteorol. Zeitschrift*, **15**, 259–263, doi:10.1127/0941-2948/2006/0130.
38 [http://www.schweizerbart.de/papers/metz/detail/15/55034/World_Map_of_the_Koppen_Geiger_climat
39 e_classificat?af=crossref](http://www.schweizerbart.de/papers/metz/detail/15/55034/World_Map_of_the_Koppen_Geiger_climate_classificat?af=crossref) (Accessed May 7, 2018).
- 40 Koven, C. D., and Coauthors, 2015: A simplified, data-constrained approach to estimate the permafrost
41 carbon-climate feedback. *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, **373**, 20140423,
42 doi:10.1098/rsta.2014.0423. <http://rsta.royalsocietypublishing.org/lookup/doi/10.1098/rsta.2014.0423>.
- 43 Kramer, K., S. J. Vreugdenhil, and D. C. van der Werf, 2008: Effects of flooding on the recruitment, damage
44 and mortality of riparian tree species: A field and simulation study on the Rhine floodplain. *For. Ecol.
45 Manage.*, **255**, 3893–3903, doi:10.1016/j.foreco.2008.03.044.
46 <https://www.sciencedirect.com/science/article/pii/S0378112708002922?via%3Dihub> (Accessed
47 October 31, 2018).
- 48 Krawchuk, M. A., and M. A. Moritz, 2011: Constraints on global fire activity vary across a resource
49 gradient. *Ecology*, **92**, 121–132, doi:10.1890/09-1843.1. <https://doi.org/10.1890/09-1843.1>.
- 50 Kreidenweis, U., and Coauthors, 2016: Afforestation to mitigate climate change: impacts on food prices
51 under consideration of albedo effects. *Environ. Res. Lett.*, **11**, 85001, doi:doi:10.1088/1748-
52 9326/11/8/085001.
- 53 Kriegler, E., and Coauthors, 2017: Fossil-fueled development (SSP5): An energy and resource intensive
54 scenario for the 21st century. *Glob. Environ. Chang.*, **42**, 297–315,
55 doi:10.1016/J.GLOENVCHA.2016.05.015.
- 56 Krishnan, R., and Coauthors, 2016: Deciphering the desiccation trend of the South Asian monsoon

- 1 hydroclimate in a warming world. *Clim. Dyn.*, **47**, 1007–1027, doi:10.1007/s00382-015-2886-5.
2 <http://link.springer.com/10.1007/s00382-015-2886-5> (Accessed March 24, 2019).
- 3 Kuang, W., T. Yang, A. Liu, C. Zhang, D. Lu, and W. Chi, 2017: An EcoCity model for regulating urban
4 land cover structure and thermal environment: Taking Beijing as an example. *Sci. China Earth Sci.*, **60**,
5 1098–1109, doi:10.1007/s11430-016-9032-9. <https://doi.org/10.1007/s11430-016-9032-9>.
- 6 Kulmala, M., 2004: A new feedback mechanism linking forests, aerosols, and climate. *Atmos. Chem. Phys.*,
7 **4**, 557–562, doi:10.5194/acp-4-557-2004.
- 8 ———, and Coauthors, 2011: General overview: European Integrated project on Aerosol Cloud Climate and
9 Air Quality interactions (EUCAARI)-integrating aerosol research from nano to global scales. *Atmos.*
10 *Chem. Phys.*, **11**, 13061–130143, doi:10.5194/acp-11-13061-2011.
- 11 Kulmala, M., T. Nieminen, R. Chellapermal, R. Makkonen, J. Bäck, and V.-M. Kerminen, 2013: Climate
12 Feedbacks Linking the Increasing Atmospheric CO₂ Concentration, BVOC Emissions, Aerosols and
13 Clouds in Forest Ecosystems. *Biology, Controls and Models of Tree Volatile Organic Compound*
14 *Emissions*, Ü. Niinemets and Monson R., Eds., Springer, Dordrecht, 489–508.
- 15 ———, and Coauthors, 2014: CO₂-induced terrestrial climate feedback mechanism: From carbon sink to
16 aerosol source and back. *Boreal Environ. Res.*, **19**, 122–131.
- 17 Kundzewicz, Z. W., and P. Germany, 2012: Changes in Impacts of Climate Extremes : Human Systems and
18 Ecosystems Coordinating Lead Authors : Lead Authors : Review Editors : Contributing Authors :
19 *Managing the Risks of Extreme Events and Disasters to Advance Climate Change Adaptation - A*
20 *Special Report of Working Groups I and II of the Intergovernmental Panel on Climate Change (IPCC)*,
21 C.B. Field et al., Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY,
22 USA, 231–290.
- 23 Kundzewicz, Z. W., and Coauthors, 2014: Flood risk and climate change: global and regional perspectives.
24 *Hydrol. Sci. J.*, **59**, 1–28, doi:10.1080/02626667.2013.857411.
25 <http://www.tandfonline.com/doi/abs/10.1080/02626667.2013.857411> (Accessed October 30, 2018).
- 26 Kunkel, K. E., and Coauthors, 2013: Monitoring and understanding trends in extreme storms: State of
27 knowledge. *Bull. Am. Meteorol. Soc.*, **94**, 499–514, doi:10.1175/BAMS-D-11-00262.1.
28 <http://journals.ametsoc.org/doi/abs/10.1175/BAMS-D-11-00262.1> (Accessed May 16, 2018).
- 29 Kurz, W. A., C. C. Dymond, G. Stinson, G. J. Rampley, E. T. Neilson, A. L. Carroll, T. Ebata, and L.
30 Safranyik, 2008a: Mountain pine beetle and forest carbon feedback to climate change. *Nature*, **452**,
31 987–990, doi:10.1038/nature06777. <http://www.ncbi.nlm.nih.gov/pubmed/18432244>.
- 32 ———, C. Smyth, and T. Lemprière, 2016: Climate change mitigation through forest sector activities:
33 principles, potential and priorities. *Unasylva*, **67**, 61–67.
- 34 Kurz, W. A. A., C. C. Dymond, G. Stinson, G. J. J. Rampley, E. T. T. Neilson, A. L. Carroll, T. Ebata, and
35 L. Safranyik, 2008b: Mountain pine beetle and forest carbon feedback to climate change. *Nature*, **452**,
36 987–990, doi:10.1038/nature06777. <http://www.ncbi.nlm.nih.gov/pubmed/18432244>.
- 37 Kuzyakov, Y., and E. Blagodatskaya, 2015: Microbial hotspots and hot moments in soil: Concept & review.
38 *Soil Biol. Biochem.*, **83**, 184–199, doi:https://doi.org/10.1016/j.soilbio.2015.01.025.
39 <http://www.sciencedirect.com/science/article/pii/S0038071715000449>.
- 40 ———, I. Bogomolova, and B. Glaser, 2014: Biochar stability in soil: Decomposition during eight years and
41 transformation as assessed by compound-specific¹⁴C analysis. *Soil Biol. Biochem.*, **70**, 229–236,
42 doi:10.1016/j.soilbio.2013.12.021.
43 <http://www.sciencedirect.com/science/article/pii/S0038071713004653>.
- 44 Kvalevåg, M. M., G. Myhre, G. Bonan, and S. Levis, 2010: Anthropogenic land cover changes in a GCM
45 with surface albedo changes based on MODIS data. *Int. J. Climatol.*, doi:10.1002/joc.2012.
- 46 van der Laan-Luijkx, I. T., and Coauthors, 2017: The CarbonTracker Data Assimilation Shell (CTDAS)
47 v1.0: Implementation and global carbon balance 2001–2015. *Geosci. Model Dev.*, **10**, 2785–2800,
48 doi:10.5194/gmd-10-2785-2017. <https://www.geosci-model-dev.net/10/2785/2017/>.
- 49 Laganière, J., D. Paré, E. Thiffault, and P. Y. Bernier, 2017: Range and uncertainties in estimating delays in
50 greenhouse gas mitigation potential of forest bioenergy sourced from Canadian forests. *GCB*
51 *Bioenergy*, **9**, 358–369, doi:10.1111/gcbb.12327.
- 52 Laguë, M. M., and A. L. S. Swann, 2016b: Progressive Midlatitude Afforestation: Impacts on Clouds, Global
53 Energy Transport, and Precipitation. *J. Clim.*, **29**, 5561–5573, doi:10.1175/JCLI-D-15-0748.1.
- 54 ———, and A. L. S. Swann, 2016a: Progressive midlatitude afforestation: Impacts on clouds, global energy
55 transport, and precipitation. *J. Clim.*, **29**, 5561–5573, doi:10.1175/jcli-d-15-0748.1.
- 56 Lai, L., C. Oh Hong, S. Kumar, S. L. Osborne, R. M. Lehman, and V. N. Owens, 2017: Soil nitrogen

- 1 dynamics in switchgrass seeded to a marginal cropland in South Dakota. *GCB Bioenergy*, n/a-n/a,
2 doi:10.1111/gcbb.12475. <http://dx.doi.org/10.1111/gcbb.12475>.
- 3 Lajtha, K., R. D. Bowden, and K. Nadelhoffer, 2014a: Litter and Root Manipulations Provide Insights into
4 Soil Organic Matter Dynamics and Stability. *Soil Sci. Soc. Am. J.*, **78**, S261,
5 doi:10.2136/sssaj2013.08.0370nafsc. <https://www.soils.org/publications/sssaj/abstracts/78/s1/S261>.
- 6 ———, K. L. Townsend, M. G. Kramer, C. Swanston, R. D. Bowden, and K. Nadelhoffer, 2014b: Changes to
7 particulate versus mineral-associated soil carbon after 50 years of litter manipulation in forest and
8 prairie experimental ecosystems. *Biogeochemistry*, **119**, 341–360, doi:10.1007/s10533-014-9970-5.
9 <https://doi.org/10.1007/s10533-014-9970-5>.
- 10 Lal, R., 2010: Managing Soils and Ecosystems for Mitigating Anthropogenic Carbon Emissions and
11 Advancing Global Food Security. *Bioscience*, doi:10.1525/bio.2010.60.9.8.
- 12 Lal, R., 2011: Sequestering carbon in soils of agro-ecosystems. *Food Policy*, **36**, S33–S39,
13 doi:10.1016/j.foodpol.2010.12.001.
- 14 Lal, R., 2013: Soil carbon management and climate change. *Carbon Manag.*, **4**, 439–462,
15 doi:10.4155/cmt.13.31. <http://www.tandfonline.com/doi/abs/10.4155/cmt.13.31>.
- 16 ———, M. Griffin, J. Apt, L. Lave, and M. G. Morgan, 2004: Managing Soil Carbon. *Science (80-.)*, **304**,
17 393 LP-393.
- 18 Lamarque, J.-F., and Coauthors, 2010: Historical (1850–2000) gridded anthropogenic and biomass burning
19 emissions of reactive gases and aerosols: methodology and application. *Atmos. Chem. Phys.*, **10**, 7017–
20 7039, doi:10.5194/acp-10-7017-2010.
- 21 Lamarque, J. F., and Coauthors, 2013: The atmospheric chemistry and climate model intercomparison
22 Project (ACCMIP): Overview and description of models, simulations and climate diagnostics. *Geosci.*
23 *Model Dev.*, **6**, 179–206, doi:10.5194/gmd-6-179-2013.
- 24 Lamers, P., and M. Junginger, 2013: The “debt” is in the detail: A synthesis of recent temporal forest carbon
25 analyses on woody biomass for energy. *Biofuels, Bioprod. Biorefining*, **7**, 373–385,
26 doi:10.1002/bbb.1407. <http://dx.doi.org/10.1002/bbb.1407>.
- 27 Landhäusser, S. M., D. Deshaies, and V. J. Lieffers, 2010: Disturbance facilitates rapid range expansion of
28 aspen into higher elevations of the Rocky Mountains under a warming climate. *J. Biogeogr.*, **37**, 68–76,
29 doi:10.1111/j.1365-2699.2009.02182.x. <http://doi.wiley.com/10.1111/j.1365-2699.2009.02182.x>
30 (Accessed May 20, 2018).
- 31 Landry, J. S., and H. D. Matthews, 2016: Non-deforestation fire vs. fossil fuel combustion: The source of
32 CO₂ emissions affects the global carbon cycle and climate responses. *Biogeosciences*, **13**, 2137–2149,
33 doi:10.5194/bg-13-2137-2016.
- 34 Larsen, M. A. D., J. H. Christensen, M. Drews, M. B. Butts, and J. C. Refsgaard, 2016: Local control on
35 precipitation in a fully coupled climate-hydrology model. *Sci. Rep.*, **6**, 22927, doi:10.1038/srep22927.
- 36 Laskin, A., J. Laskin, and S. A. Nizkorodov, 2015: Chemistry of Atmospheric Brown Carbon. *Chem. Rev.*,
37 **115**, 4335–4382, doi:10.1021/cr5006167.
- 38 Laurance, S., and Coauthors, 2018: Compositional response of Amazon forests to climate change. *Glob.*
39 *Chang. Biol.*, **25**, 39–56, doi:10.1111/gcb.14413.
40 <https://onlinelibrary.wiley.com/doi/abs/10.1111/gcb.14413> (Accessed March 12, 2019).
- 41 Law, D. J., and Coauthors, 2018: Bioclimatic Envelopes for Individual Demographic Events Driven by
42 Extremes: Plant Mortality from Drought and Warming. *Int. J. Plant Sci.*, **180**, 53–62,
43 doi:10.1086/700702. <https://www.journals.uchicago.edu/doi/10.1086/700702> (Accessed April 6, 2019).
- 44 Lawrence, D., and K. Vandecar, 2015: Effects of tropical deforestation on climate and agriculture. *Nat. Clim.*
45 *Chang.*, **5**, 27–36, doi:10.1038/nclimate2430.
- 46 Lawrence, D. M., and Coauthors, 2016: The Land Use Model Intercomparison Project (LUMIP) contribution
47 to CMIP6: Rationale and experimental design. *Geosci. Model Dev.*, **9**, 2973–2998, doi:10.5194/gmd-9-
48 2973-2016.
- 49 Lawrence, P. J., and Coauthors, 2012: Simulating the biogeochemical and biogeophysical impacts of
50 transient land cover change and wood harvest in the Community Climate System Model (CCSM4)
51 from 1850 to 2100. *J. Clim.*, doi:10.1175/JCLI-D-11-00256.1.
- 52 ———, D. M. Lawrence, and G. C. Hurtt, 2018: Attributing the Carbon Cycle Impacts of CMIP5 Historical
53 and Future Land Use and Land Cover Change in the Community Earth System Model (CESM1). *J.*
54 *Geophys. Res. Biogeosciences*, **123**, 1732–1755, doi:10.1029/2017JG004348.
- 55 Lawrimore, J. H., M. J. Menne, B. E. Gleason, C. N. Williams, D. B. Wuertz, R. S. Vose, and J. Rennie,
56 2011: An overview of the Global Historical Climatology Network monthly mean temperature data set,

- 1 version 3. *J. Geophys. Res. Atmos.*, **116**, doi:10.1029/2011JD016187.
2 <https://doi.org/10.1029/2011JD016187>.
- 3 Layton, K., and D. Ellison, 2016: Induced precipitation recycling (IPR): A proposed concept for increasing
4 precipitation through natural vegetation feedback mechanisms. *Ecol. Eng.*, **91**, 553–565,
5 doi:10.1016/j.ecoleng.2016.02.031.
- 6 Leakey, A. D. B., K. A. Bishop, and E. A. Ainsworth, 2012: A multi-biome gap in understanding of crop and
7 ecosystem responses to elevated CO₂. *Curr. Opin. Plant Biol.*, **15**, 228–236,
8 doi:10.1016/j.pbi.2012.01.009. <http://www.sciencedirect.com/science/article/pii/S1369526612000106>.
- 9 Lee, D. and Sanz, M. J., 2017: *UNFCCC Accounting for Forests: What's in and what's out of NDCs and*
10 *REDD+ - Climate and Land Use Alliance*.
- 11 Lee, X., and Coauthors, 2011: Observed increase in local cooling effect of deforestation at higher latitudes.
12 *Nature*, **479**, 384–387, doi:10.1038/nature10588.
- 13 Lehmann, J., D. Coumou, and K. Frieler, 2015: Increased record-breaking precipitation events under global
14 warming. *Clim. Change*, **132**, 501–515, doi:10.1007/s10584-015-1434-y.
15 <https://doi.org/10.1007/s10584-015-1434-y>.
- 16 Lehner, F., C. Deser, and B. M. Sanderson, 2018: Future risk of record-breaking summer temperatures and
17 its mitigation. *Clim. Change*, **146**, 363–375, doi:10.1007/s10584-016-1616-2.
18 <http://link.springer.com/10.1007/s10584-016-1616-2> (Accessed October 28, 2018).
- 19 Lejeune, Q., E. L. Davin, B. P. Guillod, and S. I. Seneviratne, 2015: Influence of Amazonian deforestation
20 on the future evolution of regional surface fluxes, circulation, surface temperature and precipitation.
21 *Clim. Dyn.*, **44**, 2769–2786, doi:10.1007/s00382-014-2203-8.
- 22 —, E. L. Davin, L. Gudmundsson, and S. I. Seneviratne, 2018: Historical deforestation increased the risk
23 of heat extremes in northern mid-latitudes 2. *Nat. Clim. Chang.*, **8**, 1–16, doi:10.1038/s41558-018-
24 0131-z. <https://doi.org/10.1038/s41558-018-0131-z>.
- 25 Lemordant, L., P. Gentine, M. Stefanon, P. Drobinski, and S. Fatichi, 2016: Modification of land-atmosphere
26 interactions by CO₂ effects: Implications for summer dryness and heat wave amplitude. *Geophys. Res.*
27 *Lett.*, **43**, 10,240–10,248, doi:10.1002/2016GL069896.
- 28 Lemprière, T. C., and Coauthors, 2013: Canadian boreal forests and climate change mitigation. *Environ.*
29 *Rev.*, **21**, 293–321, doi:10.1139/er-2013-0039.
- 30 van Lent, J., K. Hergoualc'H, and L. V. Verchot, 2015: Reviews and syntheses: Soil N₂O and NO emissions
31 from land use and land-use change in the tropics and subtropics: A meta-analysis. *Biogeosciences*, **12**,
32 7299–7313, doi:10.5194/bg-12-7299-2015.
- 33 Lenton, T. M., 2010: The potential for land-based biological CO₂ removal to lower future atmospheric CO₂
34 concentration. *Carbon Manag.*, doi:10.4155/cmt.10.12.
- 35 —, 2014: CHAPTER 3. The Global Potential for Carbon Dioxide Removal. *Geoengineering of the*
36 *Climate System*, R.M. Harrison and R.E. Hester, Eds., Royal Society of Chemistry, 52–79
37 <http://ebook.rsc.org/?DOI=10.1039/9781782621225-00052>.
- 38 Di Leo, N., F. J. Escobedo, and M. Dubbeling, 2016: The role of urban green infrastructure in mitigating
39 land surface temperature in Bobo-Dioulasso, Burkina Faso. *Environ. Dev. Sustain.*, **18**, 373–392,
40 doi:10.1007/s10668-015-9653-y. <https://doi.org/10.1007/s10668-015-9653-y>.
- 41 Leonard, M., and Coauthors, 2014: A compound event framework for understanding extreme impacts. *Wiley*
42 *Interdiscip. Rev. Clim. Chang.*, **5**, 113–128, doi:10.1002/wcc.252.
43 <http://doi.wiley.com/10.1002/wcc.252> (Accessed May 17, 2018).
- 44 Lesk, C., P. Rowhani, and N. Ramankutty, 2016a: Influence of extreme weather disasters on global crop
45 production. *Nature*, **529**, 84. <http://dx.doi.org/10.1038/nature16467>.
- 46 —, —, and —, 2016b: Influence of extreme weather disasters on global crop production. *Nature*,
47 **529**, 84–87, doi:10.1038/nature16467. <http://www.nature.com/articles/nature16467> (Accessed May 7,
48 2018).
- 49 Lewis, S. C., and D. J. Karoly, 2013: Anthropogenic contributions to Australia's record summer
50 temperatures of 2013. *Geophys. Res. Lett.*, **40**, 3708–3709, doi:10.1002/grl.50673.
51 <http://doi.wiley.com/10.1002/grl.50673> (Accessed October 28, 2018).
- 52 Lewis, S. M., and M. Kelly, 2014: Mapping the Potential for Biofuel Production on Marginal Lands:
53 Differences in Definitions, Data and Models across Scales. *Isprs Int. J. Geo-Information*, **3**, 430–459,
54 doi:10.3390/ijgi3020430.
- 55 Lhotka, O., J. Kyselý, and A. Farda, 2018: Climate change scenarios of heat waves in Central Europe and
56 their uncertainties. *Theor. Appl. Climatol.*, **131**, 1043–1054, doi:10.1007/s00704-016-2031-3.

- 1 <http://link.springer.com/10.1007/s00704-016-2031-3> (Accessed October 28, 2018).
- 2 Li, C., S. Frolking, and K. Butterbach-Bahl, 2005: Carbon sequestration in arable soils is likely to increase
3 nitrous oxide emissions, offsetting reductions in climate radiative forcing. *Clim. Change*, **72**, 321–338,
4 doi:10.1007/s10584-005-6791-5.
- 5 Li, D., and E. Bou-Zeid, 2013: Synergistic interactions between urban heat islands and heat waves: The
6 impact in cities is larger than the sum of its parts. *J. Appl. Meteorol. Climatol.*, **52**, 2051–2064,
7 doi:10.1175/JAMC-D-13-02.1.
- 8 ———, T. Sun, M. Liu, L. Yang, L. Wang, and Z. Gao, 2015a: Contrasting responses of urban and rural
9 surface energy budgets to heat waves explain synergies between urban heat islands and heat waves.
10 *Environ. Res. Lett.*, **10**, 54009, doi:10.1088/1748-9326/10/5/054009. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/10/i=5/a=054009)
11 [9326/10/i=5/a=054009](http://stacks.iop.org/1748-9326/10/i=5/a=054009).
- 12 ———, W. Liao, A. J. Rigden, X. Liu, D. Wang, S. Malyshev, and E. Shevliakova, 2019: Urban heat island:
13 Aerodynamics or imperviousness? *Sci. Adv.*, **5**, doi:10.1126/sciadv.aau4299.
14 <https://advances.sciencemag.org/content/5/4/eaau4299>.
- 15 Li, H., and Coauthors, 2018a: A new method to quantify surface urban heat island intensity. *Sci. Total*
16 *Environ.*, **624**, 262–272, doi:10.1016/j.scitotenv.2017.11.360.
17 <http://www.sciencedirect.com/science/article/pii/S0048969717334186>.
- 18 Li, W., and Coauthors, 2017a: Land-use and land-cover change carbon emissions between 1901 and 2012
19 constrained by biomass observations. *Biogeosciences*, **14**, 5053–5067, doi:10.5194/bg-14-5053-2017.
- 20 ———, and Coauthors, 2018b: Recent Changes in Global Photosynthesis and Terrestrial Ecosystem
21 Respiration Constrained From Multiple Observations. *Geophys. Res. Lett.*, **45**, 1058–1068,
22 doi:10.1002/2017GL076622.
- 23 Li, X., Y. Zhou, G. R. Asrar, M. Imhoff, and X. Li, 2017b: The surface urban heat island response to urban
24 expansion: A panel analysis for the conterminous United States. *Sci. Total Environ.*, **605–606**, 426–
25 435, doi:10.1016/j.scitotenv.2017.06.229.
26 <http://www.sciencedirect.com/science/article/pii/S0048969717316492>.
- 27 Li, Y., M. Zhao, S. Motesharrei, Q. Mu, E. Kalnay, and S. Li, 2015b: Local cooling and warming effects of
28 forests based on satellite observations. *Nat. Commun.*, **6**, 1–8, doi:10.1038/ncomms7603.
29 <http://dx.doi.org/10.1038/ncomms7603> (Accessed December 20, 2017).
- 30 ———, N. De Noblet-Ducoudré, E. L. Davin, S. Motesharrei, N. Zeng, S. Li, and E. Kalnay, 2016: The role of
31 spatial scale and background climate in the latitudinal temperature response to deforestation. *Earth*
32 *Syst. Dyn.*, **7**, 167–181, doi:10.5194/esd-7-167-2016.
- 33 Li, Z., and H. Fang, 2016: Impacts of climate change on water erosion: A review. *Earth-Science Rev.*, **163**,
34 94–117, doi:10.1016/j.earscirev.2016.10.004.
35 <https://www.sciencedirect.com/science/article/pii/S0012825216303555> (Accessed September 23,
36 2018).
- 37 Liang, J., X. Qi, L. Souza, and Y. Luo, 2016: Processes regulating progressive nitrogen limitation under
38 elevated carbon dioxide: A meta-analysis. *Biogeosciences*, **13**, 2689–2699, doi:10.5194/bg-13-2689-
39 2016.
- 40 Liang, P., and Y. Ding, 2017: The long-term variation of extreme heavy precipitation and its link to
41 urbanization effects in Shanghai during 1916–2014. *Adv. Atmos. Sci.*, **34**, 321–334,
42 doi:10.1007/s00376-016-6120-0.
- 43 Liao, W., D. Wang, X. Liu, G. Wang, and J. Zhang, 2017: Estimated influence of urbanization on surface
44 warming in Eastern China using time-varying land use data. *Int. J. Climatol.*, **37**, 3197–3208,
45 doi:10.1002/joc.4908.
- 46 Lickley, M., and S. Solomon, 2018: Drivers, timing and some impacts of global aridity change. *Environ.*
47 *Res. Lett.*, **13**, 104010, doi:10.1088/1748-9326/aae013.
- 48 Liska, A. J., and Coauthors, 2014: Biofuels from crop residue can reduce soil carbon and increase CO₂
49 emissions. *Nat. Clim. Chang.*, **4**, 398–401, doi:10.1038/nclimate2187.
- 50 Liu, B., and Coauthors, 2016a: Similar estimates of temperature impacts on global wheat yield by three
51 independent methods. *Nat. Clim. Chang.*, **6**, 1130–1136, doi:10.1038/nclimate3115.
- 52 Liu, C., C. E. Chung, F. Zhang, and Y. Yin, 2016b: The colors of biomass burning aerosols in the
53 atmosphere. *Sci. Rep.*, **6**, doi:10.1038/srep28267.
- 54 Liu, J., and H. Yang, 2010: Spatially explicit assessment of global consumptive water uses in cropland:
55 Green and blue water. *J. Hydrol.*, **384**, 187–197, doi:10.1016/j.jhydrol.2009.11.024.
56 <https://www.sciencedirect.com/science/article/pii/S0022169409007446> (Accessed March 25, 2019).

- 1 Liu, J., and Coauthors, 2017a: Contrasting carbon cycle responses of the tropical continents to the 2015–
2 2016 El Niño. *Science* (80-.), doi:10.1126/science.aam5690.
- 3 Liu, L., and T. L. Greaver, 2009: A review of nitrogen enrichment effects on three biogenic GHGs: The CO
4 2 sink may be largely offset by stimulated N₂O and CH₄ emission. *Ecol. Lett.*, **12**, 1103–1117,
5 doi:10.1111/j.1461-0248.2009.01351.x.
- 6 ———, J. S. King, F. L. Booker, C. P. Giardina, H. Lee Allen, and S. Hu, 2009: Enhanced litter input rather
7 than changes in litter chemistry drive soil carbon and nitrogen cycles under elevated CO₂: A
8 microcosm study. *Glob. Chang. Biol.*, **15**, 441–453, doi:10.1111/j.1365-2486.2008.01747.x.
9 <https://doi.org/10.1111/j.1365-2486.2008.01747.x>.
- 10 Liu, W. J., and Coauthors, 2017b: Repackaging precipitation into fewer, larger storms reduces ecosystem
11 exchanges of CO₂ and H₂O in a semiarid steppe. *Agric. For. Meteorol.*, **247**, 356–364,
12 doi:10.1016/j.agrformet.2017.08.029.
13 <https://www.sciencedirect.com/science/article/pii/S016819231730285X> (Accessed November 2, 2018).
- 14 Liu, Y., J. Stanturf, and S. Goodrick, 2010: Trends in global wildfire potential in a changing climate. *For.*
15 *Ecol. Manage.*, **259**, 685–697, doi:10.1016/j.foreco.2009.09.002.
- 16 Liu, Y., Y. Li, S. Li, and S. Motesharrei, 2015: Spatial and temporal patterns of global NDVI trends:
17 Correlations with climate and human factors. *Remote Sens.*, **7**, 13233–13250, doi:10.3390/rs71013233.
- 18 Liu, Y., and Coauthors, 2016c: Isoprene photochemistry over the Amazon rainforest. *Proc. Natl. Acad. Sci.*,
19 doi:10.1073/pnas.1524136113.
- 20 Liu, Z., 2012: Dynamics of interdecadal climate variability: A historical perspective. *J. Clim.*, **25**, 1963–
21 1995, doi:10.1175/2011JCLI3980.1.
- 22 Llonch, P., M. J. Haskell, R. J. Dewhurst, and S. P. Turner, 2017: Current available strategies to mitigate
23 greenhouse gas emissions in livestock systems: An animal welfare perspective. *Animal*,
24 doi:10.1017/S1751731116001440.
- 25 Lloyd, A. H., T. S. Rupp, C. L. Fastie, and A. M. Starfield, 2003: Patterns and dynamics of treeline advance
26 on the Seward Peninsula, Alaska. **108**, doi:10.1029/2001JD000852.
- 27 Lloyd, J., and G. D. Farquhar, 2008: Effects of rising temperatures and [CO₂] on the physiology of tropical
28 forest trees. *Philos. Trans. R. Soc. B Biol. Sci.*, **363**, 1811–1817, doi:10.1098/rstb.2007.0032.
29 <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2007.0032>.
- 30 Loarie, S. R., D. B. Lobell, G. P. Asner, Q. Mu, and C. B. Field, 2011: Direct impacts on local climate of
31 sugar-cane expansion in Brazil. *Nat. Clim. Chang.*, **1**, 105–109, doi:10.1038/nclimate1067.
- 32 Lobell, D. B., and C. Bonfils, 2008: The effect of irrigation on regional temperatures: A spatial and temporal
33 analysis of trends in California, 1934–2002. *J. Clim.*, **21**, 2063–2071, doi:10.1175/2007JCLI1755.1.
- 34 ———, and C. Tebaldi, 2014: Getting caught with our plants down: The risks of a global crop yield slowdown
35 from climate trends in the next two decades. *Environ. Res. Lett.*, **9**, 74003, doi:10.1088/1748-
36 9326/9/7/074003.
- 37 Lobell, D. B., G. Bala, and P. B. Duffy, 2006: Biogeophysical impacts of cropland management changes on
38 climate. *Geophys. Res. Lett.*, **33**, 4–7, doi:10.1029/2005GL025492.
- 39 Lobell, D. B., C. J. Bonfils, L. M. Kueppers, and M. A. Snyder, 2008: Irrigation cooling effect on
40 temperature and heat index extremes. *Geophys. Res. Lett.*, **35**, 1–5, doi:10.1029/2008GL034145.
- 41 ———, W. Schlenker, and J. Costa-Roberts, 2011: Climate trends and global crop production since 1980.
42 *Science* (80-.), **333**, 616–620, doi:10.1126/science.1204531.
43 <http://www.ncbi.nlm.nih.gov/pubmed/21551030> (Accessed December 20, 2017).
- 44 Loehman, R. A., E. Reinhardt, and K. L. Riley, 2014: Wildland fire emissions, carbon, and climate: Seeing
45 the forest and the trees - A cross-scale assessment of wildfire and carbon dynamics in fire-prone,
46 forested ecosystems. *For. Ecol. Manage.*, **317**, 9–19, doi:10.1016/j.foreco.2013.04.014.
47 <http://www.sciencedirect.com/science/article/pii/S0378112713002272>.
- 48 Lokoshchenko, M. A., 2017: Urban heat island and urban dry island in Moscow and their centennial
49 changes. *J. Appl. Meteorol. Climatol.*, **56**, 2729–2745, doi:10.1175/JAMC-D-16-0383.1.
- 50 Lombardozi, D. L., G. B. Bonan, W. Wieder, A. S. Grandy, C. Morris, and D. L. Lawrence, 2018: Cover
51 Crops May Cause Winter Warming in Snow-Covered Regions. *Geophys. Res. Lett.*, **45**, 9889–9897,
52 doi:10.1029/2018GL079000.
- 53 Longobardi, P., A. Montenegro, H. Beltrami, and M. Eby, 2016a: Deforestation induced climate change:
54 Effects of spatial scale. *PLoS One*, doi:10.1371/journal.pone.0153357.
- 55 ———, ———, ———, and ———, 2016b: Deforestation induced climate change: Effects of spatial scale. *PLoS*
56 *One*, **11**, doi:10.1371/journal.pone.0153357.

- 1 Van Loon, A. F., and G. Laaha, 2015: Hydrological drought severity explained by climate and catchment
2 characteristics. *J. Hydrol.*, **526**, 3–14, doi:10.1016/j.jhydrol.2014.10.059.
3 <https://www.sciencedirect.com/science/article/pii/S0022169414008543> (Accessed May 24, 2018).
- 4 Lopez, H., R. West, S. Dong, G. Goni, B. Kirtman, S. K. Lee, and R. Atlas, 2018: Early emergence of
5 anthropogenically forced heat waves in the western United States and Great Lakes. *Nat. Clim. Chang.*,
6 **8**, 414–420, doi:10.1038/s41558-018-0116-y. <http://www.nature.com/articles/s41558-018-0116-y>
7 (Accessed October 28, 2018).
- 8 Loranty, M. M., L. T. Berner, S. J. Goetz, Y. Jin, and J. T. Randerson, 2014: Vegetation controls on northern
9 high latitude snow-albedo feedback: Observations and CMIP5 model simulations. *Glob. Chang. Biol.*,
10 **20**, 594–606, doi:10.1111/gcb.12391.
- 11 Lorenz, R., and A. J. Pitman, 2014: Effect of land-atmosphere coupling strength on impacts from Amazonian
12 deforestation. *Geophys. Res. Lett.*, **41**, 5987–5995, doi:10.1002/2014GL061017.
13 <http://doi.wiley.com/10.1002/2014GL061017>.
- 14 —, and Coauthors, 2016: Influence of land-atmosphere feedbacks on temperature and precipitation
15 extremes in the GLACE-CMIP5 ensemble. *J. Geophys. Res.*, **121**, 607–623,
16 doi:10.1002/2015JD024053.
- 17 Di Lorenzo, E., and Coauthors, 2008: North Pacific Gyre Oscillation links ocean climate and ecosystem
18 change. *Geophys. Res. Lett.*, **35**, L08607, doi:10.1029/2007GL032838.
19 <http://doi.wiley.com/10.1029/2007GL032838> (Accessed August 29, 2018).
- 20 Los, S. O., 2013: Analysis of trends in fused AVHRR and MODIS NDVI data for 1982-2006: Indication for
21 a CO₂ fertilization effect in global vegetation. *Global Biogeochem. Cycles*, **27**, 318–330,
22 doi:10.1002/gbc.20027.
- 23 Lucht, W., and Coauthors, 1995: Climatic Control of the High-Latitude Vegetation Greening Trend and
24 Pinatubo Effect. 8–11.
- 25 Luedeling, E., 2012: Climate change impacts on winter chill for temperate fruit and nut production: A
26 review. *Sci. Hortic. (Amsterdam)*, **144**, 218–229, doi:10.1016/j.scienta.2012.07.011.
- 27 Luo, Y., S. Wan, D. Hui, and L. L. Wallace, 2001: Acclimatization of soil respiration to warming in a tall
28 grass prairie. *Nature*, **413**, 622–625, doi:10.1038/35098065. <http://dx.doi.org/10.1038/35098065>.
- 29 Luo, Z., E. Wang, and O. J. Sun, 2010: Can no-tillage stimulate carbon sequestration in agricultural soils? A
30 meta-analysis of paired experiments. *Agric. Ecosyst. Environ.*, **139**, 224–231,
31 doi:10.1016/j.agee.2010.08.006.
- 32 Lützw, M. V., I. Kögel-Knabner, K. Ekschmitt, E. Matzner, G. Guggenberger, B. Marschner, and H. Flessa,
33 2006: Stabilization of organic matter in temperate soils: Mechanisms and their relevance under
34 different soil conditions - A review. *Eur. J. Soil Sci.*, **57**, 426–445, doi:10.1111/j.1365-
35 2389.2006.00809.x. <https://doi.org/10.1111/j.1365-2389.2006.00809.x>.
- 36 Luysaert, S., and Coauthors, 2014a: Land management and land-cover change have impacts of similar
37 magnitude on surface temperature. *Nat. Clim. Chang.*, **4**, 389–393, doi:10.1038/nclimate2196.
38 <http://www.nature.com/doi/10.1038/nclimate2196> (Accessed December 20, 2017).
- 39 —, and Coauthors, 2014b: Land management and land-cover change have impacts of similar magnitude
40 on surface temperature. *Nat. Clim. Chang.*, **4**, 389–393, doi:10.1038/nclimate2196.
- 41 —, and Coauthors, 2018: Trade-offs in using European forests to meet climate objectives. *Nature*, **562**,
42 259–262, doi:10.1038/s41586-018-0577-1.
- 43 M., H. A., A. W. S., and P. T. J., 2014: The effect of urbanisation on the climatology of thunderstorm
44 initiation. *Q. J. R. Meteorol. Soc.*, **141**, 663–675, doi:10.1002/qj.2499.
- 45 Ma, D., M. Notaro, Z. Liu, G. Chen, and Y. Liu, 2013a: Simulated impacts of afforestation in East China
46 monsoon region as modulated by ocean variability. *Clim. Dyn.*, **41**, 2439–2450, doi:10.1007/s00382-
47 012-1592-9.
- 48 Ma, E., A. Liu, X. Li, F. Wu, and J. Zhan, 2013b: Impacts of vegetation change on the regional surface
49 climate: A scenario-based analysis of afforestation in Jiangxi Province, China. *Adv. Meteorol.*, **2013**,
50 doi:10.1155/2013/796163.
- 51 Mabuchi, K., Y. Sato, and H. Kida, 2005: Climatic impact of vegetation change in the Asian tropical region.
52 Part I: Case of the Northern Hemisphere summer. *J. Clim.*, **18**, 410–428, doi:10.1175/JCLI-3273.1.
- 53 MacDermott, H. J., R. J. Fensham, Q. Hua, and D. M. J. S. Bowman, 2016: Vegetation, fire and soil
54 feedbacks of dynamic boundaries between rainforest, savanna and grassland. *Austral Ecol.*, **42**, 154–
55 164, doi:10.1111/aec.12415.
- 56 MacDougall, A. H., N. C. Swart, and R. Knutti, 2016: The Uncertainty in the Transient Climate Response to

- 1 Cumulative CO₂ Emissions Arising from the Uncertainty in Physical Climate Parameters. *J. Clim.*,
2 doi:10.1175/jcli-d-16-0205.1.
- 3 Macintosh, A., H. Keith, and D. Lindenmayer, 2015: Rethinking forest carbon assessments to account for
4 policy institutions. *Nat. Clim. Chang.*, **5**, 946.
- 5 Magarini, A., and A. Calori, 2015: *Food and the Cities. Food policies for sustainable cities*. 1st ed. Edizioni
6 Ambiente, Milán,.
- 7 Maher, P., G. K. Vallis, S. C. Sherwood, M. J. Webb, and P. G. Sansom, 2018: The Impact of Parameterized
8 Convection on Climatological Precipitation in Atmospheric Global Climate Models. *Geophys. Res.
9 Lett.*, **45**, 3728–3736, doi:10.1002/2017GL076826. <http://doi.wiley.com/10.1002/2017GL076826>
10 (Accessed October 30, 2018).
- 11 Mahmood, R., S. A. Foster, T. Keeling, K. G. Hubbard, C. Carlson, and R. Leeper, 2006: Impacts of
12 irrigation on 20th century temperature in the northern Great Plains. *Glob. Planet. Change*, **54**, 1–18,
13 doi:10.1016/j.gloplacha.2005.10.004.
- 14 Mahmood, R., K. von Salzen, M. Flanner, M. Sand, J. Langner, H. Wang, and L. Huang, 2016: Seasonality
15 of global and Arctic black carbon processes in the Arctic Monitoring and Assessment Programme
16 models. *J. Geophys. Res. Atmos.*, **121**, 7100–7116, doi:10.1002/2016JD024849.
17 <http://doi.wiley.com/10.1002/2016JD024849>.
- 18 Mahony, C. R., A. J. Cannon, T. Wang, and S. N. Aitken, 2017: A closer look at novel climates: new
19 methods and insights at continental to landscape scales. *Glob. Chang. Biol.*, **23**, 3934–3955,
20 doi:10.1111/gcb.13645.
- 21 Mahowald, N. M., D. S. Ward, S. C. Doney, P. G. Hess, and J. T. Randerson, 2017: Are the impacts of land
22 use on warming underestimated in climate policy? *Environ. Res. Lett.*, **12**, 94016, doi:10.1088/1748-
23 9326/aa836d. <http://stacks.iop.org/1748-9326/12/i=9/a=094016>.
- 24 Makarieva, A. M., V. G. Gorshkov, and B. L. Li, 2009: Precipitation on land versus distance from the ocean:
25 Evidence for a forest pump of atmospheric moisture. *Ecol. Complex.*, **6**, 302–307,
26 doi:10.1016/j.ecocom.2008.11.004.
- 27 Makkonen, R., A. Asmi, V. M. Kerminen, M. Boy, A. Arneth, P. Hari, and M. Kulmala, 2012: Air pollution
28 control and decreasing new particle formation lead to strong climate warming. *Atmos. Chem. Phys.*, **12**,
29 1515–1524, doi:10.5194/acp-12-1515-2012.
- 30 Malhi, Y., C. Doughty, and D. Galbraith, 2011: The allocation of ecosystem net primary productivity in
31 tropical forests. *Philos. Trans. R. Soc. B Biol. Sci.*, **366**, 3225–3245, doi:10.1098/rstb.2011.0062.
- 32 —, and Coauthors, 2015: The linkages between photosynthesis, productivity, growth and biomass in
33 lowland Amazonian forests. *Glob. Chang. Biol.*, **21**, 2283–2295, doi:10.1111/gcb.12859.
- 34 Manabe, S., R. J. Stouffer, M. J. Spelman, and K. Bryan, 1991: Transient Responses of a Coupled Ocean–
35 Atmosphere Model to Gradual Changes of Atmospheric CO₂. Part I. Annual Mean Response. *J. Clim.*,
36 **4**, 785–818, doi:10.1175/1520-0442(1991)004<0785:TROACO>2.0.CO;2.
37 [https://doi.org/10.1175/1520-0442\(1991\)004%3C0785:TROACO%3E2.0.CO](https://doi.org/10.1175/1520-0442(1991)004%3C0785:TROACO%3E2.0.CO).
- 38 Manish, S., and Coauthors, 2017: Recent advances in understanding secondary organic aerosol: Implications
39 for global climate forcing. *Rev. Geophys.*, **55**, 509–559, doi:10.1002/2016RG000540.
- 40 Manning, D. A. C., 2008: Biological enhancement of soil carbonate precipitation: passive removal of
41 atmospheric CO₂. *Mineral. Mag.*, **72**, 639–649,
42 doi:10.1180/minmag.2008.072.2.639.
- 43 Manola, I., B. Van Den Hurk, H. De Moel, and J. C. J. H. Aerts, 2018: Future extreme precipitation
44 intensities based on a historic event. *Hydrol. Earth Syst. Sci.*, **22**, 3777–3788, doi:10.5194/hess-22-
45 3777-2018. <https://www.hydrol-earth-syst-sci.net/22/3777/2018/> (Accessed March 24, 2019).
- 46 Mao, J., and Coauthors, 2016: Human-induced greening of the northern extratropical land surface. *Nat. Clim.
47 Chang.*, **6**, 959–963, doi:10.1038/nclimate3056.
- 48 van Marle, M. J. E., and Coauthors, 2017: Fire and deforestation dynamics in Amazonia (1973–2014).
49 *Global Biogeochem. Cycles*, **31**, 24–38, doi:10.1002/2016GB005445.
- 50 Van Marle, M. J. E., and Coauthors, 2017a: Historic global biomass burning emissions for CMIP6
51 (BB4CMIP) based on merging satellite observations with proxies and fire models (1750–2015). *Geosci.
52 Model Dev.*, **10**, 3329–3357, doi:10.5194/gmd-10-3329-2017.
- 53 —, and Coauthors, 2017b: Historic global biomass burning emissions for CMIP6 (BB4CMIP) based on
54 merging satellite observations with proxies and fire models (1750–2015). *Geosci. Model Dev.*, **10**,
55 3329–3357, doi:10.5194/gmd-10-3329-2017.
- 56 Marlon, J. R., P. J. Bartlein, A. L. Daniau, S. P. Harrison, S. Y. Maezumi, M. J. Power, W. Tinner, and B.

- 1 Vanni re, 2013: Global biomass burning: A synthesis and review of Holocene paleofire records and
2 their controls. *Quat. Sci. Rev.*, **65**, 5–25, doi:10.1016/j.quascirev.2012.11.029.
- 3 Marlon, J. R., and Coauthors, 2016: Reconstructions of biomass burning from sediment-charcoal
4 \hack{\newline}records to improve data--model comparisons. *Biogeosciences*, **13**, 3225–3244,
5 doi:10.5194/bg-13-3225-2016. <https://www.biogeosciences.net/13/3225/2016/>.
- 6 Marschner, B., and Coauthors, 2008: How relevant is recalcitrance for the stabilization of organic matter in
7 soils? *J. Plant Nutr. Soil Sci.*, **171**, 91–110, doi:10.1002/jpln.200700049.
8 <https://doi.org/10.1002/jpln.200700049>.
- 9 Martin, A., and Coauthors, 2016: *Routledge Handbook of Ecosystem Services*. 1st ed. M. Potschin, R.
10 Haines-Young, R. Fish, and R.K. Turner, Eds. Routledge, Oxford, UK, 630 pp.
- 11 Masih, I., S. Maskey, F. E. F. Muss a, and P. Trambauer, 2014: A review of droughts on the African
12 continent: a geospatial and long-term perspective. *Hydrol. Earth Syst. Sci.*, **18**, 3635–3649,
13 doi:10.5194/hess-18-3635-2014. <https://www.hydrol-earth-syst-sci.net/18/3635/2014/> (Accessed
14 October 29, 2018).
- 15 Matthews, H. D., N. P. Gillett, P. A. Stott, and K. Zickfeld, 2009: The proportionality of global warming to
16 cumulative carbon emissions. *Nature*, **459**, 829. <https://doi.org/10.1038/nature08047>.
- 17 Matthews, H. D. D., A. J. J. Weaver, K. J. J. Meissner, N. P. P. Gillett, and M. Eby, 2004: Natural and
18 anthropogenic climate change: incorporating historical land cover change, vegetation dynamics and the
19 global carbon cycle. *Clim. Dyn.*, **22**, 461–479, doi:10.1007/s00382-004-0392-2.
- 20 Maule, C. F., T. Mendlik, and O. B. Christensen, 2017: The effect of the pathway to a two degrees warmer
21 world on the regional temperature change of Europe. *Clim. Serv.*, **7**, 3–11,
22 doi:10.1016/j.cliser.2016.07.002.
- 23 Mazzoncini, M., T. B. Sapkota, P. B arberi, D. Antichi, and R. Risaliti, 2011: Long-term effect of tillage,
24 nitrogen fertilization and cover crops on soil organic carbon and total nitrogen content. *Soil Tillage
25 Res.*, **114**, 165–174, doi:10.1016/j.still.2011.05.001.
- 26 McDaniel, M. D., D. Saha, M. G. Dumont, M. Hern andez, and M. A. Adams, 2019: The Effect of Land-Use
27 Change on Soil CH₄ and N₂O Fluxes: A Global Meta-Analysis. *Ecosystems*, doi:10.1007/s10021-019-
28 00347-z. <http://link.springer.com/10.1007/s10021-019-00347-z>.
- 29 McDowell, N., and Coauthors, 2018: Drivers and mechanisms of tree mortality in moist tropical forests. *New
30 Phytol.*, doi:10.1111/nph.15027.
- 31 McDowell, N. G., and C. D. Allen, 2015: Darcy’s law predicts widespread forest mortality under climate
32 warming. *Nat. Clim. Chang.*, **5**, 669–672, doi:10.1038/nclimate2641.
33 <http://www.nature.com/articles/nclimate2641> (Accessed October 25, 2018).
- 34 McDowell, N. G., D. J. Beerling, D. D. Breshears, R. A. Fisher, K. F. Raffa, and M. Stitt, 2011: The
35 interdependence of mechanisms underlying climate-driven vegetation mortality. *Trends Ecol. Evol.*, **26**,
36 523–532, doi:<https://doi.org/10.1016/j.tree.2011.06.003>.
37 <http://www.sciencedirect.com/science/article/pii/S0169534711001698>.
- 38 McGuire, A. D., and Coauthors, 2018: Dependence of the evolution of carbon dynamics in the northern
39 permafrost region on the trajectory of climate change. *Proc. Natl. Acad. Sci.*, **115**, 201719903,
40 doi:10.1073/pnas.1719903115.
- 41 McLaren, D., 2012: A comparative global assessment of potential negative emissions technologies. *Process
42 Saf. Environ. Prot.*, **90**, 489–500, doi:10.1016/j.psep.2012.10.005.
- 43 McLeod, E., and Coauthors, 2011: A blueprint for blue carbon: Toward an improved understanding of the
44 role of vegetated coastal habitats in sequestering CO₂. *Front. Ecol. Environ.*, **9**, 552–560,
45 doi:10.1890/110004. <https://doi.org/10.1890/110004>.
- 46 McLeod, J., M. Shepherd, and C. E. Konrad, 2017: Spatio-temporal rainfall patterns around Atlanta, Georgia
47 and possible relationships to urban land cover. *Urban Clim.*, **21**, 27–42,
48 doi:10.1016/j.uclim.2017.03.004.
- 49 McNorton, J., and Coauthors, 2016: Role of OH variability in the stalling of the global atmospheric CH₄
50 growth rate from 1999 to 2006. *Atmos. Chem. Phys.*, **16**, 7943–7956, doi:10.5194/acp-16-7943-2016.
- 51 Meier, M., J. Fuhrer, and A. Holzk amper, 2018: Changing risk of spring frost damage in grapevines due to
52 climate change? A case study in the Swiss Rhone Valley. *Int. J. Biometeorol.*, **62**, 991–1002,
53 doi:10.1007/s00484-018-1501-y. <http://link.springer.com/10.1007/s00484-018-1501-y> (Accessed April
54 7, 2019).
- 55 Meli, P., K. D. Holl, J. M. R. Benayas, H. P. Jones, P. C. Jones, D. Montoya, and D. M. Mateos, 2017: A
56 global review of past land use, climate, and active vs. passive restoration effects on forest recovery.

- 1 *PLoS One*, doi:10.1371/journal.pone.0171368.
- 2 Melillo, J. M., and Coauthors, 2017: Long-term pattern and magnitude of soil carbon feedback to the climate
3 system in a warming world. *Science (80-.)*, **358**, 101–105, doi:10.1126/science.aan2874.
4 <http://science.sciencemag.org/content/358/6359/101>.
- 5 Melton, J. R., and Coauthors, 2013: Geoscientific Instrumentation Methods and Data Systems Present state
6 of global wetland extent and wetland methane modelling: conclusions from a model inter-comparison
7 project (WETCHIMP). *Biogeosciences*, **10**, 753–788, doi:10.5194/bg-10-753-2013.
- 8 Menne, M. J., C. N. Williams, B. E. Gleason, J. J. Rennie, and J. H. Lawrimore, 2018: The Global Historical
9 Climatology Network Monthly Temperature Dataset, Version 4. *J. Clim.*, **31**, 9835–9854,
10 doi:10.1175/JCLI-D-18-0094.1. <https://doi.org/10.1175/JCLI-D-18-0094.1>.
- 11 Mercado, L. M., and Coauthors, 2018: Large sensitivity in land carbon storage due to geographical and
12 temporal variation in the thermal response of photosynthetic capacity. *New Phytol.*, **218**, 1462–1477,
13 doi:10.1111/nph.15100.
- 14 Meurer, K. H. E., U. Franko, C. F. Stange, J. D. Rosa, B. E. Madari, and H. F. Jungkunst, 2016: Direct
15 nitrous oxide (N₂O) fluxes from soils under different land use in Brazil—a critical review. *Environ.*
16 *Res. Lett.*, **11**, 23001, doi:10.1088/1748-9326/11/2/023001. <http://stacks.iop.org/1748-9326/11/i=2/a=023001?key=crossref.8368fef46ccb826bb457fb5e070b99b>.
- 17 Meurer, K. H. E., N. R. Haddaway, M. A. Bolinder, and T. Kätterer, 2018: Tillage intensity affects total
18 SOC stocks in boreo-temperate regions only in the topsoil—A systematic review using an ESM
19 approach. *Earth-Science Rev.*, **177**, 613–622, doi:10.1016/j.earscirev.2017.12.015.
20 <https://linkinghub.elsevier.com/retrieve/pii/S0012825217305135>.
- 21 Michalský, M., and P. S. Hooda, 2015: Greenhouse gas emissions of imported and locally produced fruit and
22 vegetable commodities: A quantitative assessment. *Environ. Sci. Policy*, **48**, 32–43,
23 doi:<https://doi.org/10.1016/j.envsci.2014.12.018>.
24 <http://www.sciencedirect.com/science/article/pii/S1462901114002469>.
- 25 Millar, C. I., and N. L. Stephenson, 2015: Temperate forest health in an era of emerging megadisturbance.
26 *Science (80-.)*, **349**, 823–826, doi:10.1126/science.aaa9933.
27 <http://www.ncbi.nlm.nih.gov/pubmed/26293954> (Accessed May 16, 2018).
- 28 Millar, R. J., and Coauthors, 2017: Emission budgets and pathways consistent with limiting warming to 1.5
29 °C. *Nat. Geosci.*, **10**, 741–747, doi:10.1038/NGEO3031.
- 30 Milly, P. C. D., and K. A. Dunne, 2016: Potential evapotranspiration and continental drying. *Nat. Clim.*
31 *Chang.*, **6**, 946–949, doi:10.1038/nclimate3046. <http://www.nature.com/articles/nclimate3046>
32 (Accessed September 22, 2018).
- 33 Min, S. K., X. Zhang, F. W. Zwiers, and G. C. Hegerl, 2011: Human contribution to more-intense
34 precipitation extremes. *Nature*, **470**, 378–381, doi:10.1038/nature09763.
- 35 Minayeva, T. Y., and A. A. Sirin, 2012: Peatland biodiversity and climate change. *Biol. Bull. Rev.*,
36 doi:10.1134/s207908641202003x.
- 37 Miner, R., 2010: *Impact of the global forest industry on atmospheric greenhouse gases*. Rome, 71 pp.
38 <http://library.wur.nl/WebQuery/clc/1941440>.
- 39 Ming, J., C. Xiao, Z. Du, and X. Yang, 2013: An overview of black carbon deposition in High Asia glaciers
40 and its impacts on radiation balance. *Adv. Water Resour.*, **55**, 80–87,
41 doi:<https://doi.org/10.1016/j.advwatres.2012.05.015>.
42 <http://www.sciencedirect.com/science/article/pii/S0309170812001601>.
- 43 Minx, J. C., and Coauthors, 2018: Negative emissions - Part 1: Research landscape and synthesis. *Environ.*
44 *Res. Lett.*, doi:10.1088/1748-9326/aabf9b.
- 45 Miralles, D. G., A. J. Teuling, C. C. Van Heerwaarden, and J. V. G. De Arellano, 2014: Mega-heatwave
46 temperatures due to combined soil desiccation and atmospheric heat accumulation. *Nat. Geosci.*, **7**,
47 345–349, doi:10.1038/ngeo2141. <http://www.nature.com/articles/ngeo2141> (Accessed August 31,
48 2018).
- 49 ———, P. Gentine, S. I. Seneviratne, and A. J. Teuling, 2018: Land-atmospheric feedbacks during droughts
50 and heatwaves: state of the science and current challenges. *Ann. N. Y. Acad. Sci.*,
51 doi:10.1111/nyas.13912.
- 52 Mishra, V., A. R. Ganguly, B. Nijssen, and D. P. Lettenmaier, 2015: Changes in observed climate extremes
53 in global urban areas. *Environ. Res. Lett.*, **10**, 24005, doi:10.1088/1748-9326/10/2/024005.
54 <http://stacks.iop.org/1748-9326/10/i=2/a=024005?key=crossref.66eab66a177894abf6a0f8b554f971af>
55 (Accessed May 17, 2018).
- 56

- 1 Mitchard, E. T. A., S. S. Saatchi, A. Baccini, G. P. Asner, S. J. Goetz, N. L. Harris, and S. Brown, 2013:
2 Uncertainty in the spatial distribution of tropical forest biomass: A comparison of pan-tropical maps.
3 *Carbon Balance Manag.*, doi:10.1186/1750-0680-8-10.
- 4 Mitchell, S. R., M. E. Harmon, and K. E. B. O'Connell, 2012: Carbon debt and carbon sequestration parity
5 in forest bioenergy production. *GCB Bioenergy*, doi:10.1111/j.1757-1707.2012.01173.x.
- 6 Mohan, T. S., and M. Rajeevan, 2017: Past and future trends of hydroclimatic intensity over the Indian
7 monsoon region. *J. Geophys. Res. Atmos.*, **122**, 896–909, doi:10.1002/2016JD025301.
8 <http://doi.wiley.com/10.1002/2016JD025301> (Accessed May 24, 2018).
- 9 Monard, C., C. Mchergui, N. Nunan, F. Martin-Laurent, and L. Vieublé-Gonod, 2012: Impact of soil matrix
10 potential on the fine-scale spatial distribution and activity of specific microbial degrader communities.
11 *FEMS Microbiol. Ecol.*, **81**, 673–683, doi:10.1111/j.1574-6941.2012.01398.x.
- 12 Mondal, N., and R. Sukumar, 2016: Fires in Seasonally Dry Tropical Forest: Testing the Varying Constraints
13 Hypothesis across a Regional Rainfall Gradient. *PLoS One*, **11**, e0159691.
- 14 Monforti, F., E. Lugato, V. Motola, K. Bodis, N. Scarlat, and J. F. Dallemand, 2015: Optimal energy use of
15 agricultural crop residues preserving soil organic carbon stocks in Europe. *Renew. Sustain. Energy*
16 *Rev.*, doi:10.1016/j.rser.2014.12.033.
- 17 Montenegro, A., M. Eby, Q. Mu, M. Mulligan, A. J. Weaver, E. C. Wiebe, and M. Zhao, 2009: The net
18 carbon drawdown of small scale afforestation from satellite observations. *Glob. Planet. Change*, **69**,
19 195–204, doi:<https://doi.org/10.1016/j.gloplacha.2009.08.005>.
20 <http://www.sciencedirect.com/science/article/pii/S0921818109001416>.
- 21 Moody, J. A., R. A. Shakesby, P. R. Robichaud, S. H. Cannon, and D. A. Martin, 2013: Current research
22 issues related to post-wildfire runoff and erosion processes. *Earth-Science Rev.*, **122**, 10–37,
23 doi:<https://doi.org/10.1016/j.earscirev.2013.03.004>.
24 <http://www.sciencedirect.com/science/article/pii/S0012825213000536>.
- 25 Moore, K. E., D. R. Fitzjarrald, R. K. Sakai, M. L. Goulden, J. W. Munger, and S. C. Wofsy, 1996: Seasonal
26 Variation in Radiative and Turbulent Exchange at a Deciduous Forest in Central Massachusetts. *J.*
27 *Appl. Meteorology*, **35**, 122–134, doi:10.1175/1520-0450(1996)035<0122:SVIRAT>2.0.CO;2.
28 [http://journals.ametsoc.org/doi/abs/10.1175/1520-
29 0450%281996%29035%3C0122%3ASVIRAT%3E2.0.CO%3B2](http://journals.ametsoc.org/doi/abs/10.1175/1520-0450%281996%29035%3C0122%3ASVIRAT%3E2.0.CO%3B2).
- 30 Mora, C., and Coauthors, 2013: The projected timing of climate departure from recent variability. *Nature*,
31 **502**, 183–187, doi:10.1038/nature12540.
- 32 —, and Coauthors, 2014: Mora et al. reply. *Nature*, **511**, E5–E6, doi:10.1038/nature13524.
- 33 —, and Coauthors, 2017: Global risk of deadly heat. *Nat. Clim. Chang.*, **7**, 501–506,
34 doi:10.1038/nclimate3322. <http://www.nature.com/doi/finder/10.1038/nclimate3322> (Accessed October
35 28, 2018).
- 36 Moragues-Faus, A., and A. Marceau, 2018: Measuring progress in sustainable food cities: An indicators
37 toolbox for action. *Sustain.*, **11**, 1–17, doi:10.3390/su11010045.
- 38 Morgan, J. A., and Coauthors, 2011: C4grasses prosper as carbon dioxide eliminates desiccation in warmed
39 semi-arid grassland. *Nature*, **476**, 202–205, doi:10.1038/nature10274.
40 <http://dx.doi.org/10.1038/nature10274>.
- 41 Moritz, E. B. and D. D. A. and M. A., 2015: A minimal model of fire-vegetation feedbacks and disturbance
42 stochasticity generates alternative stable states in grassland–shrubland–woodland systems. *Environ.*
43 *Res. Lett.*, **10**, 34018.
- 44 Moritz, M. A., and Coauthors, 2014: Learning to coexist with wildfire. *Nature*, **515**, 58.
45 <http://dx.doi.org/10.1038/nature13946>.
- 46 Mosley, L. M., 2015: Drought impacts on the water quality of freshwater systems; review and integration.
47 *Earth-Science Rev.*, **140**, 203–214, doi:10.1016/j.earscirev.2014.11.010.
48 <https://www.sciencedirect.com/science/article/pii/S0012825214002086> (Accessed May 24, 2018).
- 49 Mote, P. W., and Coauthors, 2016: Perspectives on the causes of exceptionally low 2015 snowpack in the
50 western United States. *Geophys. Res. Lett.*, **43**, 10,980–10,988, doi:10.1002/2016GL069965.
51 <http://doi.wiley.com/10.1002/2016GL069965> (Accessed October 28, 2018).
- 52 Moyano, F. E., S. Manzoni, and C. Chenu, 2013: Responses of soil heterotrophic respiration to moisture
53 availability: An exploration of processes and models. *Soil Biol. Biochem.*, **59**, 72–85,
54 doi:<https://doi.org/10.1016/j.soilbio.2013.01.002>.
55 <http://www.sciencedirect.com/science/article/pii/S0038071713000138>.
- 56 Mudryk, L. R., P. J. Kushner, C. Derksen, and C. Thackeray, 2017: Snow cover response to temperature in

- 1 observational and climate model ensembles. *Geophys. Res. Lett.*, **44**, 919–926,
2 doi:10.1002/2016GL071789.
- 3 Mueller, B., and S. I. Seneviratne, 2012: Hot days induced by precipitation deficits at the global scale. *Proc.*
4 *Natl. Acad. Sci.*, **109**, 12398–12403, doi:10.1073/pnas.1204330109.
5 <http://www.pnas.org/cgi/doi/10.1073/pnas.1204330109>.
- 6 Mueller, N. D., E. E. Butler, K. A. McKinnon, A. Rhines, M. Tingley, N. M. Holbrook, and P. Huybers,
7 2015: Cooling of US Midwest summer temperature extremes from cropland intensification. *Nat. Clim.*
8 *Chang.*, **6**, 317–322, doi:10.1038/nclimate2825. <http://dx.doi.org/10.1038/nclimate2825>.
- 9 Mukherjee, S., A. Mishra, and K. E. Trenberth, 2018: Climate Change and Drought: a Perspective on
10 Drought Indices. *Curr. Clim. Chang. Reports*, **4**, 145–163, doi:10.1007/s40641-018-0098-x.
11 <http://link.springer.com/10.1007/s40641-018-0098-x> (Accessed October 28, 2018).
- 12 Müller, C., J. Elliott, A. C. Ruane, C. Rosenzweig, S. Asseng, K. J. Boote, and J. Elliott, 2017: An AgMIP
13 framework for improved agricultural representation in integrated assessment models OPEN ACCESS
14 An AgMIP framework for improved agricultural representation in integrated assessment models.
15 *Environ. Res. Lett.*, **12**, 125003. <http://stacks.iop.org/1748-9326/12/i=12/a=125003>.
- 16 Murata, T., and N. Kawai, 2018: Degradation of the urban ecosystem function due to soil sealing:
17 involvement in the heat island phenomenon and hydrologic cycle in the Tokyo metropolitan area. *Soil*
18 *Sci. Plant Nutr.*, **64**, 145–155, doi:10.1080/00380768.2018.1439342.
19 <https://doi.org/10.1080/00380768.2018.1439342>.
- 20 Myhre, G., and Coauthors, 2013: Anthropogenic and Natural Radiative Forcing. *Climate Change 2013: The*
21 *Physical Science Basis. Contribution of Working Group 1 to the Fifth Assessment Report of the*
22 *Intergovernmental Panel on Climate Change*, T.F. Stocker et al., Eds., Cambridge University Pres,
23 Cambridge, United Kingdom and New York, NY, USA, p. 714.
- 24 Myneni, R. B., C. D. Keeling, C. J. Tucker, G. Asrar, and R. R. Nemani, 1997: Increased plant growth in the
25 northern high latitudes from 1981 to 1991. *Nature*, **386**, 698–702, doi:10.1038/386698a0.
- 26 Nabuurs, G.-J., E. J. M. M. Arets, and M.-J. Schelhaas, 2018: Understanding the implications of the EU-
27 LULUCF regulation for the wood supply from EU forests to the EU. *Carbon Balance Manag.*,
28 doi:10.1186/s13021-018-0107-3.
- 29 Nabuurs, G. J., and Coauthors, 2007: In Climate Change 2007: Mitigation. Contribution of Working Group
30 III to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change. *Forestry*, 542–
31 584.
- 32 Nabuurs, G. J., E. J. M. M. Arets, and M. J. Schelhaas, 2017: European forests show no carbon debt, only a
33 long parity effect. *For. Policy Econ.*, **75**, 120–125, doi:10.1016/j.forpol.2016.10.009.
- 34 Nainggolan, D., J. de Vente, C. Boix-Fayos, M. Termansen, K. Hubacek, and M. S. Reed, 2012:
35 Afforestation, agricultural abandonment and intensification: Competing trajectories in semi-arid
36 Mediterranean agro-ecosystems. *Agric. Ecosyst. Environ.*, **159**, 90–104,
37 doi:10.1016/j.agee.2012.06.023.
38 <https://www.sciencedirect.com/science/article/pii/S0167880912002502> (Accessed September 24,
39 2018).
- 40 Naudts, K., Y. Chen, M. J. McGrath, J. Ryder, A. Valade, J. Otto, and S. Luyssaert, 2016a: Mitigate Climate
41 Warming. *Science (80-)*, **351**, 597–601, doi:10.1126/science.aac9976.
- 42 —, —, M. J. McGrath, J. Ryder, A. Valade, J. Otto, and S. Luyssaert, 2016b: Europe’s forest
43 management did not mitigate climate warming. *Science (80-)*, **351**, 597 LP-600,
44 doi:10.1126/science.aad7270.
- 45 Nazemi, A., and H. S. Wheatler, 2015: On inclusion of water resource management in Earth system models -
46 Part 1: Problem definition and representation of water demand. *Hydrol. Earth Syst. Sci.*, **19**, 33–61,
47 doi:10.5194/hess-19-33-2015.
- 48 Nemet, G. F., and Coauthors, 2018: Negative emissions - Part 3: Innovation and upscaling. *Environ. Res.*
49 *Lett.*, doi:10.1088/1748-9326/aabff4.
- 50 Neuvonen, S., P. Niemela, and T. Virtanen, 1999: Climatic Change and Insect Outbreaks in Boreal Forests:
51 The Role of Winter Temperatures. *Ecol. Bull.*, 63–67.
52 https://www.jstor.org/stable/20113228?seq=1#metadata_info_tab_contents (Accessed April 2, 2019).
- 53 Newman, L., C. Ling, and K. Peters, 2012: Between field and table: environmental implications of local food
54 distribution. *Int. J. Sustain. Soc.*, **5**, 11, doi:10.1504/ijssoc.2013.050532.
- 55 Newman, M., and Coauthors, 2016: The Pacific decadal oscillation, revisited. *J. Clim.*, **29**, 4399–4427,
56 doi:10.1175/JCLI-D-15-0508.1.

- 1 Nicely, J. M., T. P. Canty, M. Manyin, L. D. Oman, R. J. Salawitch, S. D. Steenrod, S. E. Strahan, and S. A.
2 Strode, 2018: Changes in Global Tropospheric OH Expected as a Result of Climate Change Over the
3 Last Several Decades. *J. Geophys. Res. Atmos.*, **123**, 10,774–10,795, doi:10.1029/2018JD028388.
4 <http://doi.wiley.com/10.1029/2018JD028388>.
- 5 Nicholls, N., and S. I. Seneviratne, 2015: Comparing IPCC assessments: how do the AR4 and SREX
6 assessments of changes in extremes differ? *Clim. Change*, **133**, 7–21, doi:10.1007/s10584-013-0818-0.
7 <http://link.springer.com/10.1007/s10584-013-0818-0> (Accessed October 16, 2018).
- 8 Nisbet, E. G., and Coauthors, 2019: Very strong atmospheric methane growth in the four years 2014–2017:
9 Implications for the Paris Agreement. *Global Biogeochem. Cycles*, 2018GB006009,
10 doi:10.1029/2018GB006009.
- 11 Nishina, K., and Coauthors, 2014: Quantifying uncertainties in soil carbon responses to changes in global
12 mean temperature and precipitation. *Earth Syst. Dyn.*, **5**, 197–209, doi:10.5194/esd-5-197-2014.
- 13 Niu, S., Y. Luo, D. Li, S. Cao, J. Xia, J. Li, and M. D. Smith, 2014: Plant growth and mortality under
14 climatic extremes: An overview. *Environ. Exp. Bot.*, **98**, 13–19, doi:10.1016/j.envexpbot.2013.10.004.
15 <https://www-sciencedirect-com.ezproxy.uct.ac.za/science/article/pii/S0098847213001482> (Accessed
16 May 25, 2018).
- 17 Niyogi, D., C. Kishtawal, S. Tripathi, and R. S. Govindaraju, 2010: Observational evidence that agricultural
18 intensification and land use change may be reducing the Indian summer monsoon rainfall. *Water*
19 *Resour. Res.*, **46**, 1–17, doi:10.1029/2008WR007082.
- 20 Nizeyimana, E. L., G. W. Petersen, M. L. Imhoff, H. R. Sinclair, S. W. Waltman, D. S. Reed-Margetan, E.
21 R. Levine, and J. M. Russo, 2001: Assessing the Impact of Land Conversion to Urban Use on Soils
22 with Different Productivity Levels in the USA. *Soil Sci. Soc. Am. J.*, **65**, 391–402,
23 doi:10.2136/sssaj2001.652391x. <http://dx.doi.org/10.2136/sssaj2001.652391x>.
- 24 de Noblet-Ducoudré, N., M. Claussen, and C. Prentice, 2000: Mid-Holocene greening of the Sahara: First
25 results of the GAIM 6000 year BP experiment with two asynchronously coupled atmosphere/biome
26 models. *Clim. Dyn.*, **16**, 643–659, doi:10.1007/s003820000074.
- 27 De Noblet-Ducoudré, N., and Coauthors, 2012: Determining robust impacts of land-use-induced land cover
28 changes on surface climate over North America and Eurasia: Results from the first set of LUCID
29 experiments. *J. Clim.*, **25**, 3261–3281, doi:10.1175/JCLI-D-11-00338.1.
- 30 de Noblet, N. I., I. Colin Prentice, S. Joussaume, D. Texier, A. Botta, and A. Haxeltine, 1996: Possible role
31 of atmosphere-biosphere interactions in triggering the last glaciation. *Geophys. Res. Lett.*, **23**, 3191–
32 3194, doi:10.1029/96GL03004.
- 33 Nogherotto, R., E. Coppola, F. Giorgi, and L. Mariotti, 2013: Impact of Congo Basin deforestation on the
34 African monsoon. *Atmos. Sci. Lett.*, **14**, 45–51, doi:10.1002/asl2.416.
- 35 Noormets, A., D. Epron, J. C. Domec, S. G. McNulty, T. Fox, G. Sun, and J. S. King, 2015: Effects of forest
36 management on productivity and carbon sequestration: A review and hypothesis. *For. Ecol. Manage.*,
37 **355**, 124–140, doi:https://doi.org/10.1016/j.foreco.2015.05.019.
- 38 Norby, R. J., J. M. Warren, C. M. Iversen, B. E. Medlyn, and R. E. McMurtrie, 2010: CO₂ enhancement of
39 forest productivity constrained by limited nitrogen availability. *Proc. Natl. Acad. Sci.*, **107**, 19368–
40 19373, doi:10.1073/pnas.1006463107. <http://www.pnas.org/cgi/doi/10.1073/pnas.1006463107>.
- 41 Nousiainen, T., 2011: Optical modeling of mineral dust particles : A review. *J. Quant. Spectrosc. Radiat.*
42 *Transf.*, 5–6.
- 43 Novick, K. A., and Coauthors, 2016: The increasing importance of atmospheric demand for ecosystem water
44 and carbon fluxes. *Nat. Clim. Chang.*, **6**, 1023–1027, doi:10.1038/nclimate3114.
- 45 O’Gorman, P. A., 2014: Contrasting responses of mean and extreme snowfall to climate change. *Nature*,
46 **512**, 416–418, doi:10.1038/nature13625. <http://www.nature.com/articles/nature13625> (Accessed May
47 16, 2018).
- 48 O’Gorman, P. A., 2015: Precipitation Extremes Under Climate Change. *Curr. Clim. Chang. Reports*, **1**, 49–
49 59, doi:10.1007/s40641-015-0009-3. <http://link.springer.com/10.1007/s40641-015-0009-3> (Accessed
50 May 19, 2018).
- 51 O’Halloran, T. L., and Coauthors, 2012: Radiative forcing of natural forest disturbances. *Glob. Chang. Biol.*,
52 **18**, 555–565, doi:10.1111/j.1365-2486.2011.02577.x.
- 53 O’ishi, R., and A. Abe-Ouchi, 2009: Influence of dynamic vegetation on climate change arising from
54 increasing CO₂. *Clim. Dyn.*, **33**, 645–663, doi:10.1007/s00382-009-0611-y.
- 55 Oates, L. G., D. S. Duncan, I. Gelfand, N. Millar, G. P. Robertson, and R. D. Jackson, 2016: Nitrous oxide
56 emissions during establishment of eight alternative cellulosic bioenergy cropping systems in the North

- 1 Central United States. *GCB Bioenergy*, **8**, 539–549, doi:10.1111/gcbb.12268.
2 <http://dx.doi.org/10.1111/gcbb.12268>.
- 3 Obersteiner, M., and Coauthors, 2016: Assessing the land resource–food price nexus of the Sustainable
4 Development Goals. *Sci. Adv.*, **2**, e1501499–e1501499, doi:10.1126/sciadv.1501499.
- 5 Oenema, O., and Coauthors, 2014: Reducing nitrous oxide emissions from the global food system. *Curr.*
6 *Opin. Environ. Sustain.*, **9–10**, 55–64, doi:10.1016/j.cosust.2014.08.003.
7 <http://dx.doi.org/10.1016/j.cosust.2014.08.003>.
- 8 Oke, T. R., G. Mills, A. Christen, and J. A. Voogt, 2017: *Urban Climates*. Cambridge University Press,
9 Cambridge,.
- 10 Okuno, M., and T. Nakamura, 2003: Radiocarbon dating of tephra layers: recent progress in Japan. *Quat.*
11 *Int.*, **105**, 49–56, doi:https://doi.org/10.1016/S1040-6182(02)00150-7.
12 <http://www.sciencedirect.com/science/article/pii/S1040618202001507>.
- 13 Olefeldt, D., and Coauthors, 2016: Circumpolar distribution and carbon storage of thermokarst landscapes.
14 *Nat. Commun.*, **7**, 1–11, doi:10.1038/ncomms13043.
- 15 Oliveira, P. H. F., and Coauthors, 2007: The effects of biomass burning aerosols and clouds on the CO₂ flux
16 in Amazonia. *Tellus, Ser. B Chem. Phys. Meteorol.*, **59**, 338–349, doi:10.1111/j.1600-
17 0889.2007.00270.x.
- 18 De Oliveira Bordonal, R., R. Lal, D. Alves Aguiar, E. B. De Figueiredo, L. Ito Perillo, M. Adami, B. F.
19 Theodor Rudorff, and N. La Scala, 2015: Greenhouse gas balance from cultivation and direct land use
20 change of recently established sugarcane (*Saccharum officinarum*) plantation in south-central Brazil.
21 *Renew. Sustain. Energy Rev.*, doi:10.1016/j.rser.2015.07.137.
- 22 Ometto, J. P. H. B., A. D. Nobre, H. R. Rocha, P. Artaxo, and L. A. Martinelli, 2005: Amazonia and the
23 modern carbon cycle: Lessons learned. *Oecologia*, **143**, 483–500, doi:10.1007/s00442-005-0034-3.
- 24 Oort, A. H., and J. P. Peixóto, 1983: Global angular momentum and energy balance requirements from
25 observations. *Adv. Geophys.*, **25**, 355–490, doi:10.1016/S0065-2687(08)60177-6.
26 <https://www.sciencedirect.com/science/article/pii/S0065268708601776> (Accessed May 7, 2018).
- 27 Van Oost, K., and Coauthors, 2007: The Impact of Agricultural Soil Erosion on the Global Carbon Cycle.
28 *Science (80-.)*, **318**, 626 LP-629.
- 29 Orlowsky, B., and S. I. Seneviratne, 2012: Global changes in extreme events: Regional and seasonal
30 dimension. *Clim. Change*, **110**, 669–696, doi:10.1007/s10584-011-0122-9.
- 31 Ornstein, L., I. Aleinov, and D. Rind, 2009: Irrigated afforestation of the Sahara and Australian Outback to
32 end global warming. *Clim. Change*, **97**, 409–437, doi:10.1007/s10584-009-9626-y.
- 33 Orwin, K. H., M. U. F. Kirschbaum, M. G. St John, and I. A. Dickie, 2011: Organic nutrient uptake by
34 mycorrhizal fungi enhances ecosystem carbon storage: A model-based assessment. *Ecol. Lett.*, **14**, 493–
35 502, doi:10.1111/j.1461-0248.2011.01611.x. <https://doi.org/10.1111/j.1461-0248.2011.01611.x>.
- 36 Osterloh, K., N. Tauchnitz, O. Spott, J. Hepp, S. Bernsdorf, and R. Meissner, 2018: Changes of methane and
37 nitrous oxide emissions in a transition bog in central Germany (German National Park Harz Mountains)
38 after rewetting. *Wetl. Ecol. Manag.*, **26**, 87–102, doi:10.1007/s11273-017-9555-x.
39 <http://link.springer.com/10.1007/s11273-017-9555-x>.
- 40 Otto, F. E. L., N. Massey, G. J. Van Oldenborgh, R. G. Jones, and M. R. Allen, 2012: Reconciling two
41 approaches to attribution of the 2010 Russian heat wave. *Geophys. Res. Lett.*, **39**, n/a–n/a,
42 doi:10.1029/2011GL050422. <http://doi.wiley.com/10.1029/2011GL050422> (Accessed October 30,
43 2018).
- 44 Paasonen, P., and Coauthors, 2013: Warming-induced increase in aerosol number concentration likely to
45 moderate climate change. *Nat. Geosci.*, **6**, 438–442, doi:10.1038/ngeo1800.
- 46 Pacifico, F., G. A. Folberth, C. D. Jones, S. P. Harrison, and W. J. Collins, 2012: Sensitivity of biogenic
47 isoprene emissions to past, present, and future environmental conditions and implications for
48 atmospheric chemistry. *J. Geophys. Res. Atmos.*, **117**, doi:10.1029/2012JD018276.
- 49 Pal, J. S., and E. A. B. Eltahir, 2016: Future temperature in southwest Asia projected to exceed a threshold
50 for human adaptability. *Nat. Clim. Chang.*, **6**, 197–200, doi:10.1038/nclimate2833.
51 <http://www.nature.com/articles/nclimate2833> (Accessed October 28, 2018).
- 52 Pall, P., M. R. Allen, and D. A. Stone, 2007: Testing the Clausius-Clapeyron constraint on changes in
53 extreme precipitation under CO₂ warming. *Clim. Dyn.*, **28**, 351–363, doi:10.1007/s00382-006-0180-2.
54 <http://link.springer.com/10.1007/s00382-006-0180-2> (Accessed May 19, 2018).
- 55 Pall, P., T. Aina, D. A. Stone, P. A. Stott, T. Nozawa, A. G. J. Hilberts, D. Lohmann, and M. R. Allen, 2011:
56 Anthropogenic greenhouse gas contribution to flood risk in England and Wales in autumn 2000.

- 1 *Nature*, **470**, 382–385, doi:10.1038/nature09762. <http://www.nature.com/articles/nature09762>
2 (Accessed May 25, 2018).
- 3 Palm, C., H. Blanco-Canqui, F. DeClerck, L. Gatere, and P. Grace, 2014: Conservation agriculture and
4 ecosystem services: An overview. *Agric. Ecosyst. Environ.*, **187**, 87–105,
5 doi:10.1016/j.agee.2013.10.010.
- 6 Pan, N., X. Feng, B. Fu, S. Wang, F. Ji, and S. Pan, 2018: Increasing global vegetation browning hidden in
7 overall vegetation greening: Insights from time-varying trends. *Remote Sens. Environ.*, **214**, 59–72,
8 doi:10.1016/j.rse.2018.05.018.
9 <https://www.sciencedirect.com/science/article/abs/pii/S0034425718302396> (Accessed October 16,
10 2018).
- 11 Pan, Y., and Coauthors, 2011a: A large and persistent carbon sink in the world's forests. *Science (80-.)*,
12 **333**, 988–993, doi:10.1126/science.1201609. <http://www.ncbi.nlm.nih.gov/pubmed/21764754>
13 (Accessed October 31, 2018).
- 14 —, and Coauthors, 2011b: A large and persistent carbon sink in the world's forests. *Science (80-.)*, **333**,
15 988–993, doi:10.1126/science.1201609.
- 16 Pandey, B., Q. Zhang, and K. C. Seto, 2018: Time series analysis of satellite data to characterize multiple
17 land use transitions: a case study of urban growth and agricultural land loss in India. *J. Land Use Sci.*,
18 **13**, 221–237, doi:10.1080/1747423X.2018.1533042. <https://doi.org/10.1080/1747423X.2018.1533042>.
- 19 Pangala, S. R., and Coauthors, 2017: Large emissions from floodplain trees close the Amazon methane
20 budget. *Nature*, **552**, 230–234, doi:10.1038/nature24639. <http://dx.doi.org/10.1038/nature24639>.
- 21 Park, S., and Coauthors, 2012: Trends and seasonal cycles in the isotopic composition of nitrous oxide since
22 1940. *Nat. Geosci.*, **5**, 261–265, doi:10.1038/ngeo1421. <http://dx.doi.org/10.1038/ngeo1421>.
- 23 Parmesan, C., and G. Yohe, 2003: A globally coherent fingerprint of climate change impacts across natural
24 systems. *Nature*, **421**, 37–42, doi:10.1038/nature01286. <http://www.nature.com/articles/nature01286>
25 (Accessed May 7, 2018).
- 26 Parry, M., C. Rosenzweig, and M. Livermore, 2005: Climate change, global food supply and risk of hunger.
27 *Philos. Trans. R. Soc. B Biol. Sci.*, **360**, 2125–2138, doi:10.1098/rstb.2005.1751.
28 <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2005.1751>.
- 29 Parry, M. L., C. Rosenzweig, A. Iglesias, M. Livermore, and G. Fischer, 2004: Effects of climate change on
30 global food production under SRES emissions and socio-economic scenarios. *Glob. Environ. Chang.*,
31 **14**, 53–67, doi:10.1016/j.gloenvcha.2003.10.008.
- 32 Paschalis, A., G. G. Katul, S. Fatichi, S. Palmroth, and D. Way, 2017: On the variability of the ecosystem
33 response to elevated atmospheric CO₂ across spatial and temporal scales at the Duke Forest FACE
34 experiment. *Agric. For. Meteorol.*, **232**, 367–383, doi:10.1016/j.agrformet.2016.09.003.
35 <http://dx.doi.org/10.1016/j.agrformet.2016.09.003>.
- 36 Páscoa, P., C. M. Gouveia, A. Russo, and R. M. Trigo, 2017: Drought trends in the Iberian Peninsula over
37 the last 112 years. *Adv. Meteorol.*, **2017**, 1–13, doi:10.1155/2017/4653126.
38 <https://www.hindawi.com/journals/amete/2017/4653126/> (Accessed October 29, 2018).
- 39 Pataki, D. E., C. G. Boone, T. S. Hogue, G. D. Jenerette, J. P. McFadden, and S. Pincetl, 2011: Socio-
40 ecohydrology and the urban water challenge. *Ecohydrology*, **4**, 341–347, doi:10.1002/eco.209.
41 <https://doi.org/10.1002/eco.209>.
- 42 Patra, P. K., and Coauthors, 2016: Regional Methane Emission Estimation Based on Observed Atmospheric
43 Concentrations (2002–2012). *J. Meteorol. Soc. Japan. Ser. II*, **94**, 91–113, doi:10.2151/jmsj.2016-006.
44 https://www.jstage.jst.go.jp/article/jmsj/94/1/94_2016-006/_article.
- 45 Pau, S., D. K. Okamoto, O. Calderón, and S. J. Wright, 2018: Long-term increases in tropical flowering
46 activity across growth forms in response to rising CO₂ and climate change. *Glob. Chang. Biol.*, **24**,
47 2105–2116, doi:10.1111/gcb.14004. <http://doi.wiley.com/10.1111/gcb.14004> (Accessed January 8,
48 2018).
- 49 Pausas, J. G., and S. Paula, 2012: Fuel shapes the fire-climate relationship: Evidence from Mediterranean
50 ecosystems. *Glob. Ecol. Biogeogr.*, **21**, 1074–1082, doi:10.1111/j.1466-8238.2012.00769.x.
51 <https://doi.org/10.1111/j.1466-8238.2012.00769.x>.
- 52 Pausas, J. G., and M. M. Millán, 2018: Greening and Browning in a Climate Change Hotspot: The
53 Mediterranean Basin. *Bioscience*, **69**, 143–151, doi:10.1093/biosci/biy157.
54 <https://academic.oup.com/bioscience/article/69/2/143/5254231> (Accessed March 13, 2019).
- 55 Paustian, K., J. Six, E. T. Elliott, and H. W. Hunt, 2000: Management options for reducing CO₂ emissions
56 from agricultural soils. *Biogeochemistry*, **48**, 147–163, doi:10.1023/A:1006271331703.

- 1 Paustian, K., J. Lehmann, S. Ogle, D. Reay, G. P. Robertson, and P. Smith, 2016: Climate-smart soils.
2 *Nature*, **532**, 49–57, doi:10.1038/nature17174.
- 3 Pawson, S. M., A. Brin, E. G. Brockerhoff, D. Lamb, T. W. Payn, A. Paquette, and J. A. Parrotta, 2013:
4 Plantation forests, climate change and biodiversity. *Biodivers. Conserv.*, **22**, 1203–1227,
5 doi:10.1007/s10531-013-0458-8. <http://link.springer.com/10.1007/s10531-013-0458-8> (Accessed
6 October 31, 2018).
- 7 Pechony, O., and D. T. Shindell, 2010: Driving forces of global wildfires over the past millennium and the
8 forthcoming century. *Proc. Natl. Acad. Sci.*, **107**, 19167–19170, doi:10.1073/pnas.1003669107.
9 <http://www.pnas.org/cgi/doi/10.1073/pnas.1003669107>.
- 10 Pecl, G. T., and Coauthors, 2017: Biodiversity redistribution under climate change: Impacts on ecosystems
11 and human well-being. *Science (80-.)*, **355**, eaai9214, doi:10.1126/science.aai9214.
- 12 Pellitier, P. T., and D. R. Zak, 2018: Ectomycorrhizal fungi and the enzymatic liberation of nitrogen from
13 soil organic matter: why evolutionary history matters. *New Phytol.*, **217**, 68–73,
14 doi:10.1111/nph.14598. <https://doi.org/10.1111/nph.14598>.
- 15 Peñaloza, D., M. Erlandsson, J. Berlin, M. Wålander, and A. Falk, 2018: Future scenarios for climate
16 mitigation of new construction in Sweden: Effects of different technological pathways. *J. Clean. Prod.*,
17 doi:10.1016/j.jclepro.2018.03.285.
- 18 Pendleton, L., and Coauthors, 2012: Estimating Global “Blue Carbon” Emissions from Conversion and
19 Degradation of Vegetated Coastal Ecosystems. *PLoS One*, **7**, e43542,
20 doi:10.1371/journal.pone.0043542.
- 21 Peng, S., P. Ciais, F. Maignan, W. Li, J. Chang, T. Wang, and C. Yue, 2017: Sensitivity of land use change
22 emission estimates to historical land use and land cover mapping. *Global Biogeochem. Cycles*, **31**,
23 626–643, doi:10.1002/2015GB005360.
- 24 Peng, S.-S., and Coauthors, 2014: Afforestation in China cools local land surface temperature. *Proc. Natl.*
25 *Acad. Sci.*, **111**, 2915–2919, doi:10.1073/pnas.1315126111.
- 26 Peñuelas, J., 2009: Phenology feedbacks on climate change. *Science (80-.)*, **324**, 887–888,
27 doi:10.1126/science.1173004.
28 [isi:000266048800020%5Cnhttp://www.sciencemag.org/cgi/reprint/324/5929/887.pdf](http://www.sciencemag.org/cgi/reprint/324/5929/887.pdf)
29 <http://www.sciencemag.org/content/324/5929/887.full.pdf>.
- 30 —, and J. Llusà, 2003: BVOCs: Plant defense against climate warming? *Trends Plant Sci.*, **8**, 105–109,
31 doi:10.1016/S1360-1385(03)00008-6.
- 32 —, and M. Staudt, 2010: BVOCs and global change. *Trends Plant Sci.*, **15**, 133–144,
33 doi:10.1016/j.tplants.2009.12.005.
- 34 —, and Coauthors, 2013: Human-induced nitrogen-phosphorus imbalances alter natural and managed
35 ecosystems across the globe. *Nat. Commun.*, **4**, doi:10.1038/ncomms3934.
- 36 —, and Coauthors, 2017: Shifting from a fertilization-dominated to a warming-dominated period. *Nat.*
37 *Ecol. Evol.*, **1**, 1438–1445, doi:10.1038/s41559-017-0274-8. [http://dx.doi.org/10.1038/s41559-017-](http://dx.doi.org/10.1038/s41559-017-0274-8)
38 [0274-8](http://dx.doi.org/10.1038/s41559-017-0274-8).
- 39 Perket, J., M. G. Flanner, and J. E. Kay, 2014: Diagnosing shortwave cryosphere radiative effect and its 21st
40 century evolution in CESM. *J. Geophys. Res.*, **119**, 1356–1362, doi:10.1002/2013JD021139.
41 <https://doi.org/10.1002/2013JD021139>.
- 42 Perlwitz, J. P., C. Pérez García-Pando, and R. L. Miller, 2015: Predicting the mineral composition of dust
43 aerosols - Part 2: Model evaluation and identification of key processes with observations. *Atmos. Chem.*
44 *Phys.*, **15**, 11629–11652, doi:10.5194/acp-15-11629-2015.
- 45 Perugini, L., L. Caporaso, S. Marconi, A. Cescatti, B. Quesada, N. De Noblet-Ducoudré, J. I. House, and A.
46 Arneth, 2017: Biophysical effects on temperature and precipitation due to land cover change. *Environ.*
47 *Res. Lett.*, **12**, doi:10.1088/1748-9326/aa6b3f.
- 48 Peters, G. P., and O. Geden, 2017: Catalysing a political shift from low to negative carbon. *Nat. Clim.*
49 *Chang.*, **7**, 619–621, doi:10.1038/nclimate3369.
- 50 Peylin, P., and Coauthors, 2013: Global atmospheric carbon budget: results from an ensemble of atmospheric
51 CO₂ and CH₄ inversions. *Biogeosciences*, **10**, 6699–6720,
52 doi:10.5194/bg-10-6699-2013. <https://www.biogeosciences.net/10/6699/2013/>.
- 53 Pfahl, S., P. A. O’Gorman, and E. M. Fischer, 2017: Understanding the regional pattern of projected future
54 changes in extreme precipitation. *Nat. Clim. Chang.*, **7**, 423–427, doi:10.1038/nclimate3287.
55 <http://www.nature.com/doi/10.1038/nclimate3287> (Accessed May 21, 2018).
- 56 Philben, M., J. Holmquist, G. Macdonald, D. Duan, K. Kaiser, and R. Benner, 2015: Temperature, oxygen,

- 1 and vegetation controls on decomposition in a James Bay peatland. *Global Biogeochem. Cycles*, **29**,
2 729–743, doi:10.1002/2014GB004989. <http://doi.wiley.com/10.1002/2014GB004989> (Accessed
3 November 2, 2018).
- 4 Phillips, O. L., and Coauthors, 2009: Drought Sensitivity of the Amazon Rainforest. *Science* (80-.), **323**,
5 1344 LP-1347. <http://science.sciencemag.org/content/323/5919/1344.abstract>.
- 6 Phillips, R. P., and Coauthors, 2011: Enhanced root exudation induces microbial feedbacks to N cycling in a
7 pine forest under long-term CO₂ fumigation. *Ecol. Lett.*, **14**, 187–194, doi:10.1111/j.1461-
8 0248.2010.01570.x. <https://doi.org/10.1111/j.1461-0248.2010.01570.x>.
- 9 Phillips, T. J., M. Wickett, A. Mirin, D. B. Lobell, K. Caldeira, C. Delire, and G. Bala, 2007: Combined
10 climate and carbon-cycle effects of large-scale deforestation. *Proc. Natl. Acad. Sci.*,
11 doi:10.1073/pnas.0608998104.
- 12 Piao, S., and Coauthors, 2015: Detection and attribution of vegetation greening trend in China over the last
13 30 years. *Glob. Chang. Biol.*, **21**, 1601–1609, doi:10.1111/gcb.12795.
14 <http://doi.wiley.com/10.1111/gcb.12795> (Accessed October 31, 2018).
- 15 —, and Coauthors, 2018: Lower land-use emissions responsible for increased net land carbon sink during
16 the slow warming period. *Nat. Geosci.*, **11**, 739–743, doi:10.1038/s41561-018-0204-7.
17 <http://www.nature.com/articles/s41561-018-0204-7>.
- 18 Pielke, R. A., and Coauthors, 2011: Land use/land cover changes and climate: Modeling analysis and
19 observational evidence. *Wiley Interdiscip. Rev. Clim. Chang.*, **2**, 828–850, doi:10.1002/wcc.144.
- 20 Pilaš, I., K.-H. Feger, U. Vilhar, and A. Wahren, 2011: Multidimensionality of Scales and Approaches for
21 Forest–Water Interactions. *Forest Management and the Water Cycle*, M. Bredemeier, S. Cohen, D.L.
22 Goldbold, E. Lode, V. Pichler, and P. Schleppei, Eds., Springer, Heidelberg, 351–380
23 http://link.springer.com/10.1007/978-90-481-9834-4_20 (Accessed March 27, 2019).
- 24 Pingoud, K., T. Ekholm, R. Sievänen, S. Huuskonen, and J. Hynynen, 2018: Trade-offs between forest
25 carbon stocks and harvests in a steady state – A multi-criteria analysis. *J. Environ. Manage.*,
26 doi:10.1016/j.jenvman.2017.12.076.
- 27 Pison, I., B. Ringeval, P. Bousquet, C. Prigent, and F. Papa, 2013: Stable atmospheric methane in the 2000s:
28 Key-role of emissions from natural wetlands. *Atmos. Chem. Phys.*, **13**, 11609–11623, doi:10.5194/acp-
29 13-11609-2013.
- 30 Pitman, A., and N. de Noblet-Ducoudré, 2012: Human Effects on Climate Through Land-Use-Induced Land-
31 Cover Change. *Futur. World's Clim.*, 77–95, doi:10.1016/B978-0-12-386917-3.00004-X.
- 32 Pitman, A. J., 2003: The evolution of, and revolution in, land surface schemes designed for climate models.
33 *Int. J. Climatol.*, **23**, 479–510, doi:10.1002/joc.893.
- 34 —, F. B. Avila, G. Abramowitz, Y. P. Wang, S. J. Phipps, and N. De Noblet-Ducoudré, 2011: Importance
35 of background climate in determining impact of land-cover change on regional climate. *Nat. Clim.*
36 *Chang.*, **1**, 472–475, doi:10.1038/nclimate1294. <http://dx.doi.org/10.1038/nclimate1294>.
- 37 Pitman, A. J., and Coauthors, 2012: Effects of land cover change on temperature and rainfall extremes in
38 multi-model ensemble simulations. *Earth Syst. Dyn.*, **3**, 213–231, doi:10.5194/esd-3-213-2012.
39 <https://www.earth-syst-dynam.net/3/213/2012/>.
- 40 Plevin, R. J., J. Beckman, A. A. Golub, J. Witcover, and M. O'Hare, 2015: Carbon accounting and economic
41 model uncertainty of emissions from biofuels-induced land use change. *Environ. Sci. Technol.*,
42 doi:10.1021/es505481d.
- 43 Poeplau, C., and A. Don, 2015: Carbon sequestration in agricultural soils via cultivation of cover crops - A
44 meta-analysis. *Agric. Ecosyst. Environ.*, **200**, 33–41, doi:10.1016/j.agee.2014.10.024.
- 45 Pöhlker, M. L., and Coauthors, 2016: Long-term observations of cloud condensation nuclei in the Amazon
46 rain forest – Part 1: Aerosol size distribution, hygroscopicity, and new model parametrizations for CCN
47 prediction. *Atmos. Chem. Phys.*, **16**, 15709–15740, doi:10.5194/acp-16-15709-2016.
48 <https://www.atmos-chem-phys.net/16/15709/2016/>.
- 49 —, and Coauthors, 2018: Long-term observations of cloud condensation nuclei over the Amazon rain
50 forest - Part 2: Variability and characteristics of biomass burning, long-range transport, and pristine
51 rain forest aerosols. *Atmos. Chem. Phys.*, **18**, 10289–10331, doi:10.5194/acp-18-10289-2018.
- 52 Polydoros, A., T. Mavrakou, and C. Cartalis, 2018: Quantifying the Trends in Land Surface Temperature and
53 Surface Urban Heat Island Intensity in Mediterranean Cities in View of Smart Urbanization. *Urban*
54 *Sci.*, **2**, 16, doi:10.3390/urbansci2010016. <http://www.mdpi.com/2413-8851/2/1/16>.
- 55 Pongratz, J., C. H. Reick, T. Raddatz, and M. Claussen, 2010: Biogeophysical versus biogeochemical
56 climate response to historical anthropogenic land cover change. *Geophys. Res. Lett.*, **37**, 1–5,

- 1 doi:10.1029/2010GL043010.
- 2 ———, ———, R. A. Houghton, and J. I. House, 2014: Terminology as a key uncertainty in net land use and
3 land cover change carbon flux estimates. *Earth Syst. Dyn.*, **5**, 177–195, doi:10.5194/esd-5-177-2014.
- 4 Pongratz, J., and Coauthors, 2018: *Models meet data: Challenges and opportunities in implementing land*
5 *management in Earth system models*.
- 6 Ponomarev, I. E., I. V. Kharuk, and J. K. Ranson, 2016: Wildfires Dynamics in Siberian Larch Forests. *For.*
7 *J.*, **7**, doi:10.3390/f7060125.
- 8 Poorter, L., and Coauthors, 2016: Biomass resilience of Neotropical secondary forests. *Nature*,
9 doi:10.1038/nature16512.
- 10 Popp, A., and Coauthors, 2014: Land-use transition for bioenergy and climate stabilization: Model
11 comparison of drivers, impacts and interactions with other land use based mitigation options. *Clim.*
12 *Change*, **123**, 495–509, doi:10.1007/s10584-013-0926-x.
- 13 ———, and Coauthors, 2017: Land-use futures in the shared socio-economic pathways. *Glob. Environ.*
14 *Chang.*, **42**, 331–345, doi:10.1016/J.GLOENVCHA.2016.10.002.
- 15 Port, U., V. Brovkin, and M. Claussen, 2012: The influence of vegetation dynamics on anthropogenic
16 climate change. *Earth Syst. Dyn.*, **3**, 233–243, doi:10.5194/esd-3-233-2012.
- 17 Porter, J. R., L. Xie, A. J. Challinor, K. Cochrane, S. M. Howden, M. M. Iqbal, D. B. Lobell, and M. I.
18 Travasso, 2014: Food security and food production systems. *Climate Change 2014: Impacts,*
19 *Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II*
20 *to the Fifth Assessment Report of the Intergovernmental Panel of Climate Change*, C.B. Field et al.,
21 Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 485–533.
- 22 Porter, S. D., D. S. Reay, P. Higgins, and E. Bomberg, 2016: A half-century of production-phase greenhouse
23 gas emissions from food loss & waste in the global food supply chain. *Sci. Total Environ.*, **571**, 721–
24 729, doi:10.1016/j.scitotenv.2016.07.041.
- 25 Pöschl, U., and M. Shiraiwa, 2015: Multiphase Chemistry at the Atmosphere–Biosphere Interface
26 Influencing Climate and Public Health in the Anthropocene. *Chem. Rev.*, **115**, 4440–4475,
27 doi:10.1021/cr500487s.
- 28 Pöschl, U., and Coauthors, 2010: Rainforest aerosols as biogenic nuclei of clouds and precipitation in the
29 Amazon. *Science (80-.)*, **329**, 1513–1516, doi:10.1126/science.1191056.
30 <http://www.sciencemag.org/cgi/doi/10.1126/science.1191056>.
- 31 Posthumus, H., J. Morris, T. M. Hess, D. Neville, E. Phillips, and A. Baylis, 2009: Impacts of the summer
32 2007 floods on agriculture in England. *J. Flood Risk Manag.*, **2**, 182–189, doi:10.1111/j.1753-
33 318X.2009.01031.x. <http://doi.wiley.com/10.1111/j.1753-318X.2009.01031.x> (Accessed October 31,
34 2018).
- 35 Potter, G. L., H. W. Ellsaesser, M. C. MacCracken, and J. S. Ellis, 1981: Albedo change by man: test of
36 climatic effects. *Nature*, **291**, 47–49, doi:10.1038/291047a0. <https://doi.org/10.1038/291047a0>.
- 37 POTTER, G. L., H. W. ELLSAESSER, M. C. MACCRACKEN, and F. M. LUTHER, 1975: Possible
38 climatic impact of tropical deforestation. *Nature*, **258**, 697–698, doi:10.1038/258697a0.
39 <https://doi.org/10.1038/258697a0>.
- 40 Poulter, B., and Coauthors, 2017: Global wetland contribution to 2000–2012 atmospheric methane growth
41 rate dynamics. *Environ. Res. Lett.*, **12**, doi:10.1088/1748-9326/aa8391.
- 42 Powell, T. L., and Coauthors, 2013: Confronting model predictions of carbon fluxes with measurements of
43 Amazon forests subjected to experimental drought. *New Phytol.*, **200**, 350–365,
44 doi:10.1111/nph.12390. <https://doi.org/10.1111/nph.12390>.
- 45 Powell, T. W. R., and T. M. Lenton, 2012a: Future carbon dioxide removal via biomass energy constrained
46 by agricultural efficiency and dietary trends. *Energy Environ. Sci.*,
47 ———, and ———, 2012b: Future carbon dioxide removal via biomass energy constrained by agricultural
48 efficiency and dietary trends. *Energy Environ. Sci.*, doi:10.1039/c2ee21592f.
- 49 Powlson, D. S., A. P. Whitmore, and K. W. T. Goulding, 2011: Soil carbon sequestration to mitigate climate
50 change: A critical re-examination to identify the true and the false. *Eur. J. Soil Sci.*, **62**, 42–55,
51 doi:10.1111/j.1365-2389.2010.01342.x.
- 52 Powlson, D. S., and Coauthors, 2014a: Limited potential of no-till agriculture for climate change mitigation.
53 *Nat. Clim. Chang.*, **4**, 678–683, doi:10.1038/nclimate2292.
- 54 ———, C. M. Stirling, M. L. Jat, B. G. Gerard, C. a. Palm, P. a. Sanchez, and K. G. Cassman, 2014b: Limited
55 potential of no-till agriculture for climate change mitigation. *Nat. Clim. Chang.*, **4**, 678–683,
56 doi:10.1038/nclimate2292.

- 1 ———, ———, C. Thierfelder, R. P. White, and M. L. Jat, 2016: Does conservation agriculture deliver climate
2 change mitigation through soil carbon sequestration in tropical agro-ecosystems? *Agric. Ecosyst.*
3 *Environ.*, **220**, 164–174, doi:10.1016/j.agee.2016.01.005.
- 4 Pradhan, P., D. E. Reusser, and J. P. Kropp, 2013: Embodied Greenhouse Gas Emissions in Diets. *PLoS*
5 *One*, **8**, e62228, doi:10.1371/journal.pone.0062228.
- 6 ———, M. K. B. Lüdeke, D. E. Reusser, and J. P. Kropp, 2014: Food self-sufficiency across scales: How local
7 can we go? *Environ. Sci. Technol.*, **48**, 9463–9470, doi:10.1021/es5005939.
- 8 Prather, M. J., and Coauthors, 2017: Global atmospheric chemistry - Which air matters. *Atmos. Chem. Phys.*,
9 **17**, 9081–9102, doi:10.5194/acp-17-9081-2017.
- 10 Pratt, K., and D. Moran, 2010: Evaluating the cost-effectiveness of global biochar mitigation potential.
11 *Biomass and Bioenergy*, doi:10.1016/j.biombioe.2010.03.004.
- 12 Prein, A. F., and Coauthors, 2015: A review on regional convection-permitting climate modeling:
13 Demonstrations, prospects, and challenges. *Rev. Geophys.*, **53**, 323–361, doi:10.1002/2014RG000475.
- 14 ———, R. M. Rasmussen, K. Ikeda, C. Liu, M. P. Clark, and G. J. Holland, 2017: The future intensification of
15 hourly precipitation extremes. *Nat. Clim. Chang.*, **7**, 48–52, doi:10.1038/nclimate3168.
16 <http://www.nature.com/articles/nclimate3168> (Accessed May 24, 2018).
- 17 Prentice, I.C., Liang, X., Medlyn, B.E. and Wang, Y. P., 2015: Reliable, robust and realistic: the three R's of
18 next-generation land-surface modelling. *Atmos. Chem. Phys.*, **15**, 5987–6005.
- 19 Prevedello, J. A., G. R. Winck, M. M. Weber, E. Nichols, and B. Sinervo, 2019: Impacts of forestation and
20 deforestation on local temperature across the globe. *PLoS One*, **14**, 1–18,
21 doi:10.1371/journal.pone.0213368.
- 22 Price, H. C., and Coauthors, 2018: Atmospheric Ice-Nucleating Particles in the Dusty Tropical Atlantic. *J.*
23 *Geophys. Res. Atmos.*, **123**, 2175–2193, doi:10.1002/2017JD027560.
24 <https://doi.org/10.1002/2017JD027560>.
- 25 Prokopiou, M., C. J. Sapart, J. Rosen, P. Sperlich, T. Blunier, E. Brook, R. S. W. van de Wal, and T.
26 Röckmann, 2018: Changes in the Isotopic Signature of Atmospheric Nitrous Oxide and Its Global
27 Average Source During the Last Three Millennia. *J. Geophys. Res. Atmos.*, **123**, 10,757–10,773,
28 doi:10.1029/2018JD029008.
- 29 Prospero, J. M., P. Ginoux, O. Torres, S. E. Nicholson, and T. E. Gill, 2002: ENVIRONMENTAL
30 CHARACTERIZATION OF GLOBAL SOURCES OF ATMOSPHERIC SOIL DUST IDENTIFIED
31 WITH THE NIMBUS 7 TOTAL OZONE MAPPING SPECTROMETER (TOMS) ABSORBING
32 AEROSOL PRODUCT. *Rev. Geophys.*, **40**, 2–31, doi:10.1029/2000RG000095.
- 33 Prudhomme, C., and Coauthors, 2014: Hydrological droughts in the 21st century, hotspots and uncertainties
34 from a global multimodel ensemble experiment. *Proc. Natl. Acad. Sci.*, **111**, 3262–3267,
35 doi:10.1073/pnas.1222473110. <http://www.pnas.org/lookup/doi/10.1073/pnas.1222473110> (Accessed
36 May 24, 2018).
- 37 Pu, B., and P. Ginoux, 2017: Projection of American dustiness in the late 21st century due to climate change.
38 *Sci. Rep.*, **7**, 5553, doi:10.1038/s41598-017-05431-9.
- 39 ———, and ———, 2018: How reliable are CMIP5 models in simulating dust optical depth? *Atmos. Chem.*
40 *Phys.*, **18**, 12491–12510, doi:10.5194/acp-18-12491-2018. [https://www.atmos-chem-](https://www.atmos-chem-phys.net/18/12491/2018/)
41 [phys.net/18/12491/2018/](https://www.atmos-chem-phys.net/18/12491/2018/).
- 42 Pugh, T. A. M., and Coauthors, 2015: Simulated carbon emissions from land-use change are substantially
43 enhanced by accounting for agricultural management. *Environ. Res. Lett.*, **10**, 124008,
44 doi:10.1088/1748-9326/10/12/124008.
- 45 Pugh, T. A. M., C. Müller, J. Elliott, D. Deryng, C. Folberth, S. Olin, E. Schmid, and A. Arneth, 2016:
46 Climate analogues suggest limited potential for intensification of production on current croplands under
47 climate change. *Nat. Commun.*, **7**, 1–8, doi:10.1038/ncomms12608.
- 48 Pulliainen, J., and Coauthors, 2017: Early snowmelt significantly enhances boreal springtime carbon uptake.
49 *Proc. Natl. Acad. Sci.*, **114**, 201707889, doi:10.1073/pnas.1707889114.
- 50 Qian, H., R. Joseph, and N. Zeng, 2010: Enhanced terrestrial carbon uptake in the northern high latitudes in
51 the 21st century from the Coupled Carbon Cycle Climate Model Intercomparison Project model
52 projections. *Glob. Chang. Biol.*, **16**, 641–656, doi:10.1111/j.1365-2486.2009.01989.x.
- 53 Qin, D., M. Tignor, S. K. Allen, J. Boschung, A. Nauels, Y. Xia, V. Bex, and U. Kingdom, 2013: Annex III:
54 Glossary. *Change 2013: The Physical Science Basis. Contribution of Working Group I to the Fifth*
55 *Assessment Report of the Intergovernmental Panel on Climate Change*, T.F. Stocker et al., Eds.,
56 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 1447–1466.

- 1 Qu, X., and A. Hall, 2014: On the persistent spread in snow-albedo feedback. *Clim. Dyn.*, **42**, 69–81,
2 doi:10.1007/s00382-013-1774-0.
- 3 Quéré, C., and Coauthors, 2018: Global Carbon Budget 2018. *Earth Syst. Sci. Data*, **10**, 2141–2194,
4 doi:10.5194/essd-10-2141-2018. <https://www.earth-syst-sci-data.net/10/405/2018/>.
- 5 Le Quéré, C., and Coauthors, 2015: Global Carbon Budget 2015. *Earth Syst. Sci. Data*, **7**, 349–396,
6 doi:10.5194/essd-7-349-2015.
- 7 Le Quéré, C., and Coauthors, 2017: Global Carbon Budget 2017. *Earth Syst. Sci. Data Discuss.*, **10**, 1–79,
8 doi:10.5194/essd-2017-123. <https://www.earth-syst-sci-data-discuss.net/essd-2017-123/>.
- 9 Quesada, B., R. Vautard, P. Yiou, M. Hirschi, and S. I. Seneviratne, 2012: Asymmetric European summer
10 heat predictability from wet and dry southern winters and springs. *Nat. Clim. Chang.*, **2**, 736–741,
11 doi:10.1038/nclimate1536.
- 12 ———, A. Arneth, and N. De Noblet-Ducoudré, 2017a: Atmospheric, radiative, and hydrologic effects of
13 future land use and land cover changes: A global and multimodel climate picture. *J. Geophys. Res.*,
14 **122**, 5113–5131, doi:10.1002/2016JD025448.
- 15 ———, N. Devaraju, N. de Noblet-Ducoudré, and A. Arneth, 2017b: Reduction of monsoon rainfall in
16 response to past and future land use and land cover changes. *Geophys. Res. Lett.*, **44**, 1041–1050,
17 doi:10.1002/2016GL070663.
- 18 ———, A. Arneth, E. Robertson, and N. de Noblet-Ducoudré, 2018: Potential strong contribution of future
19 anthropogenic land-use and land-cover change to the terrestrial carbon cycle. *Environ. Res. Lett.*, **13**,
20 64023, doi:10.1088/1748-9326/aac4c3.
- 21 Rachmayani, R., M. Prange, and M. Schulz, 2015: North African vegetation-precipitation feedback in early
22 and mid-Holocene climate simulations with CCSM3-DGVM. *Clim. Past*, **11**, 175–185, doi:10.5194/cp-
23 11-175-2015.
- 24 Rafael, S., H. Martins, M. Marta-Almeida, E. Siqueira, S. Coelho, A. Rocha, C. Borrego, and M. Lopes, 2017:
25 Quantification and mapping of urban fluxes under climate change: Application of WRF-SUEWS model
26 to Greater Porto area (Portugal). *Environ. Res.*, **155**, 321–334, doi:10.1016/j.envres.2017.02.033.
- 27 Rajagopal, D., and R. J. Plevin, 2013: Implications of market-mediated emissions and uncertainty for biofuel
28 policies. *Energy Policy*, **56**, 75–82, doi:https://doi.org/10.1016/j.enpol.2012.09.076.
29 <http://www.sciencedirect.com/science/article/pii/S0301421512008646>.
- 30 Ramarao, M. V. S., R. Krishnan, J. Sanjay, and T. P. Sabin, 2015: Understanding land surface response to
31 changing South Asian monsoon in a warming climate. *Earth Syst. Dyn.*, **6**, 569–582, doi:10.5194/esd-
32 6-569-2015. <https://www.earth-syst-dynam.net/6/569/2015/> (Accessed March 20, 2019).
- 33 Ramarao, M. V. S., J. Sanjay, and R. Krishnan, 2016: Modulation of summer monsoon sub-seasonal surface
34 air temperature over India by soil moisture-temperature coupling. *Mausam*, **67**, 53–66.
- 35 Randerson, J. ., G. R. van der Werf, L. . Giglio, G. J. Collatz, and P. S. Kasibhatla, 2015: *Global Fire*
36 *Emissions Data Base, version 4*. Oak Ridge, TN, TN.
- 37 Rap, A., and Coauthors, 2018: Enhanced global primary production by biogenic aerosol via diffuse radiation
38 fertilization. *Nat. Geosci.*, doi:10.1038/s41561-018-0208-3.
- 39 Ray, D. K., J. S. Gerber, G. K. Macdonald, and P. C. West, 2015: Climate variation explains a third of global
40 crop yield variability. *Nat. Commun.*, **6**, 5989, doi:10.1038/ncomms6989.
- 41 Reddington, C. L., D. V. Spracklen, P. Artaxo, D. A. Ridley, L. V. Rizzo, and A. Arana, 2016: Analysis of
42 particulate emissions from tropical biomass burning using a global aerosol model and long-term surface
43 observations. *Atmos. Chem. Phys.*, doi:10.5194/acp-16-11083-2016.
- 44 Reed, S. C., X. Yang, and P. E. Thornton, 2015: Incorporating phosphorus cycling into global modeling
45 efforts: A worthwhile, tractable endeavor. *New Phytol.*, **208**, 324–329, doi:10.1111/nph.13521.
- 46 Reich, P. B., and S. E. Hobbie, 2013: Decade-long soil nitrogen constraint on the CO₂ fertilization of plant
47 biomass. *Nat. Clim. Chang.*, **3**, 278–282, doi:10.1038/nclimate1694.
48 <http://dx.doi.org/10.1038/nclimate1694>.
- 49 ———, ———, and T. D. Lee, 2014: Plant growth enhancement by elevated CO₂ eliminated by joint
50 water and nitrogen limitation. *Nat. Geosci.*, **7**, 920–924, doi:10.1038/ngeo2284.
- 51 ———, ———, ———, and M. A. Pastore, 2018: Unexpected reversal of C₃ versus C₄ grass response to elevated
52 CO₂ during a 20-year field experiment. *Science (80-.)*, **360**, 317–320, doi:10.1126/science.aas9313.
53 <http://www.sciencemag.org/lookup/doi/10.1126/science.aas9313>.
- 54 Reichstein, M., and Coauthors, 2013: Climate extremes and the carbon cycle. *Nature*, **500**, 287–295,
55 doi:10.1038/nature12350. <http://www.nature.com/articles/nature12350> (Accessed October 30, 2018).
- 56 Renaud, V., and M. Rebetez, 2008: Comparison between open site and below canopy climatic conditions in

- 1 {Switzerland} during the exceptionally hot summer of 2003. *Agric. For. Meteorol.*, **149**, 873–880,
2 doi:<https://doi.org/10.1016/j.agrformet.2008.11.006>.
- 3 Renaud, V., J. L. Innes, M. Dobbertin, and M. Rebetez, 2011: Comparison between open-site and below-
4 canopy climatic conditions in Switzerland for different types of forests over 10 years (1998–2007).
5 *Theor. Appl. Climatol.*, **105**, 119–127, doi:10.1007/s00704-010-0361-0.
- 6 Resplandy, L., and Coauthors, 2018: Revision of global carbon fluxes based on a reassessment of oceanic
7 and riverine carbon transport. *Nat. Geosci.*, **11**, 504–509, doi:10.1038/s41561-018-0151-3.
8 <http://www.nature.com/articles/s41561-018-0151-3>.
- 9 Revadekar, J. V., Y. K. Tiwari, and K. R. Kumar, 2012: Impact of climate variability on NDVI over the
10 Indian region during 1981–2010. *Int. J. Remote Sens.*, **33**, 7132–7150,
11 doi:10.1080/01431161.2012.697642.
- 12 Revi, A., D. E. Satterthwaite, F. Aragón-Durand, J. Corfee-Morlot, R. B. R. Kiunsi, M. Pelling, D. C.
13 Roberts, and W. Solecki, 2014: Urban areas. *Climate Change 2014: Impacts, Adaptation, and*
14 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth*
15 *Assessment Report of the Intergovernmental Panel of Climate Change*, C.B. Field et al., Eds.,
16 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 535–612.
- 17 Riahi, K., and Coauthors, 2017: The Shared Socioeconomic Pathways and their energy, land use, and
18 greenhouse gas emissions implications: An overview. *Glob. Environ. Chang.*, **42**,
19 doi:10.1016/j.gloenvcha.2016.05.009.
- 20 Rice, A. L., C. L. Butenhoff, D. G. Teama, F. H. Röger, M. A. K. Khalil, and R. A. Rasmussen, 2016:
21 Atmospheric methane isotopic record favors fossil sources flat in 1980s and 1990s with recent increase.
22 *Proc. Natl. Acad. Sci.*, **113**, 10791–10796, doi:10.1073/pnas.1522923113.
- 23 Rich, S. M., and M. Watt, 2013: Soil conditions and cereal root system architecture: Review and
24 considerations for linking Darwin and Weaver. *J. Exp. Bot.*, **64**, 1193–1208, doi:10.1093/jxb/ert043.
25 <https://academic.oup.com/jxb/article-lookup/doi/10.1093/jxb/ert043> (Accessed November 2, 2018).
- 26 Richardson, A. ., T. . Keenana, M. Migliavaccab, Y. Ryua, O. Sonnentaga, and M. Toomeya, 2013: Climate
27 change, phenology, and phenological control of vegetatif feedbacks to the climate system. *Agric. For,*
28 *Meteorol.*, **169**, 156–157, doi:<https://doi.org/10.1016/j.agrformet.2012.09.012>.
29 <http://www.sciencedirect.com/science/article/pii/S0168192312002869>.
- 30 Ridgwell, A., J. S. Singarayer, A. M. Hetherington, and P. J. Valdes, 2009: Tackling Regional Climate
31 Change By Leaf Albedo Bio-geoengineering. *Curr. Biol.*, **19**, 146–150, doi:10.1016/j.cub.2008.12.025.
32 <http://dx.doi.org/10.1016/j.cub.2008.12.025>.
- 33 Rigby, M., and Coauthors, 2013: Re-evaluation of the lifetimes of the major CFCs and Biogeosciences CH 3
34 CCl 3 using atmospheric trends. *Atmos. Chem. Phys.*, **13**, 2691–2702, doi:10.5194/acp-13-2691-2013.
- 35 Riley, W., J. Tang, and W. J. Riley, 2014: Weaker soil carbon – climate feedbacks resulting from microbial
36 and abiotic interactions Weaker soil carbon – climate feedbacks resulting from microbial and abiotic
37 interactions. *Nat. Clim. Chang.*, **5**, 56, doi:10.1038/nclimate2438.
- 38 Rivas-Martinez, S., S. Rivas-Saenz, and A. Marino, 2011: Worldwide Bioclimatic Classification System.
39 *Glob. Geobot.*, **1**, 1–638, doi:10.5616/gg110001.
- 40 Rivera Ferre, M. G., 2014: Impacts of Climate Change on Food Availability: Distribution and Exchange of
41 Food. *Global Environmental Change*, Springer Netherlands, Dordrecht, 701–707
42 http://link.springer.com/10.1007/978-94-007-5784-4_119 (Accessed September 29, 2018).
- 43 Rizzo, L. V., and Coauthors, 2013: Long term measurements of aerosol optical properties at a primary forest
44 site in Amazonia. *Atmos. Chem. Phys.*, **13**, 2391–2413, doi:10.5194/acp-13-2391-2013.
45 <https://www.atmos-chem-phys.net/13/2391/2013/>.
- 46 Robaa, S. M., 2013: Some aspects of the urban climates of Greater Cairo Region, Egypt. *Int. J. Climatol.*, **33**,
47 3206–3216, doi:10.1002/joc.3661.
- 48 Roberts, K. G., B. A. Gloy, S. Joseph, N. R. Scott, and J. Lehmann, 2010: Life cycle assessment of biochar
49 systems: Estimating the energetic, economic, and climate change potential. *Environ. Sci. Technol.*,
50 doi:10.1021/es902266r.
- 51 Roberts, P. D., G. B. Stewart, and A. S. Pullin, 2006: *Are review articles a reliable source of evidence to*
52 *support conservation and environmental management? A comparison with medicine.* 409–423 pp.
- 53 Robertson, G. P., and Coauthors, 2017: Cellulosic biofuel contributions to a sustainable energy future:
54 Choices and outcomes. *Science (80-.)*, **356**, doi:10.1126/science.aal2324.
- 55 Robinson, A. L., and Coauthors, 2007: Rethinking organic aerosols: Semivolatile emissions and
56 photochemical aging. *Science (80-.)*, **315**, 1259–1262, doi:10.1126/science.1133061.

- 1 <http://science.sciencemag.org/content/315/5816/1259.abstract>.
- 2 Rocha-Lima, A., and Coauthors, 2018: A detailed characterization of the Saharan dust collected during the
3 Fennec campaign in 2011: In situ ground-based and laboratory measurements. *Atmos. Chem. Phys.*, **18**,
4 1023–1043, doi:10.5194/acp-18-1023-2018. <https://www.atmos-chem-phys.net/18/1023/2018/>.
- 5 Roe, S., S. Weiner, M. Obersteiner, and S. Frank, 2017: *How improved land use can contribute to the 1.5C*
6 *goal of the Paris Agreement*.
- 7 Rogelj, J., and Coauthors, 2014: Disentangling the effects of CO₂ and short-lived climate forcer mitigation.
8 *Proc. Natl. Acad. Sci.*, **111**, 16325–16330, doi:10.1073/pnas.1415631111.
9 <http://www.pnas.org/lookup/doi/10.1073/pnas.1415631111>.
- 10 ———, and Coauthors, 2016: Paris Agreement climate proposals need a boost to keep warming well below
11 2 °C. *Nature*, **534**, 631. <http://dx.doi.org/10.1038/nature18307>.
- 12 ———, and Coauthors, 2018: Scenarios towards limiting global mean temperature increase below 1.5 °c. *Nat.*
13 *Clim. Chang.*, **8**, 325–332, doi:10.1038/s41558-018-0091-3.
- 14 Rogers, A., and Coauthors, 2017: A roadmap for improving the representation of photosynthesis in Earth
15 system models. *New Phytol.*, **213**, 22–42, doi:10.1111/nph.14283.
- 16 Rohde, R., and Coauthors, 2013: Berkeley earth temperature averaging process. *Geoinformatics*
17 *Geostatistics An Overv.*, **1**, 1–13.
- 18 Rosenkranz, M., T. A. M. Pugh, J.-P. Sghnitzler, and A. Arneth, 2015: Effect of land-use change and
19 management on biogenic volatile organic compound emissions - selecting climate-smart cultivars.
20 *Plant. Cell Environ.*, **38**, 1896–1912, doi:10.1111/pce.12453.
- 21 Rosenzweig, C., and Coauthors, 2013: The Agricultural Model Intercomparison and Improvement Project
22 (AgMIP): Protocols and pilot studies. *Agric. For. Meteorol.*, **170**, 166–182,
23 doi:10.1016/j.agrformet.2012.09.011.
- 24 Rossi, S., F. N. Tubiello, P. Prosperi, M. Salvatore, H. Jacobs, R. Biancalani, J. I. House, and L. Boschetti,
25 2016: FAOSTAT estimates of greenhouse gas emissions from biomass and peat fires. *Clim. Change*,
26 **135**, 699–711, doi:10.1007/s10584-015-1584-y.
- 27 Rossow, W. B., A. Mekonnen, C. Pearl, and W. Goncalves, 2013: Tropical precipitation extremes. *J. Clim.*,
28 **26**, 1457–1466, doi:10.1175/JCLI-D-11-00725.1. [http://journals.ametsoc.org/doi/abs/10.1175/JCLI-D-](http://journals.ametsoc.org/doi/abs/10.1175/JCLI-D-11-00725.1)
29 [11-00725.1](http://journals.ametsoc.org/doi/abs/10.1175/JCLI-D-11-00725.1) (Accessed October 30, 2018).
- 30 de Rouw, A., and Coauthors, 2010: Possibilities of carbon and nitrogen sequestration under conventional
31 tillage and no-till cover crop farming (Mekong valley, Laos). *Agric. Ecosyst. Environ.*,
32 doi:10.1016/j.agee.2009.12.013.
- 33 Rowe, R. L., A. M. Keith, D. Elias, M. Dondini, P. Smith, J. Oxley, and N. P. McNamara, 2016: Initial soil
34 C and land-use history determine soil C sequestration under perennial bioenergy crops. *GCB*
35 *Bioenergy*, **8**, 1046–1060, doi:10.1111/gcbb.12311. <http://dx.doi.org/10.1111/gcbb.12311>.
- 36 Rowell, D. P., and R. Chadwick, 2018: Causes of the uncertainty in projections of tropical terrestrial rainfall
37 change: East Africa. *J. Clim.*, **31**, 5977–5995, doi:10.1175/JCLI-D-17-0830.1.
38 <http://journals.ametsoc.org/doi/10.1175/JCLI-D-17-0830.1> (Accessed October 30, 2018).
- 39 Rowland, L., and Coauthors, 2015: Death from drought in tropical forests is triggered by hydraulics not
40 carbon starvation. *Nature*, **528**, 119. <https://doi.org/10.1038/nature15539>.
- 41 Roy, J., and Coauthors, 2016a: Elevated CO₂ maintains grassland net carbon uptake under a future heat and
42 drought extreme. *Proc. Natl. Acad. Sci.*, **113**, 6224–6229, doi:10.1073/pnas.1524527113.
43 <http://www.pnas.org/lookup/doi/10.1073/pnas.1524527113> (Accessed August 31, 2018).
- 44 ———, and Coauthors, 2016b: Elevated CO₂ maintains grassland net carbon uptake under a future heat and
45 drought extreme. *Proc. Natl. Acad. Sci.*, **113**, 6224–6229, doi:10.1073/pnas.1524527113.
46 <http://www.pnas.org/lookup/doi/10.1073/pnas.1524527113> (Accessed August 31, 2018).
- 47 Roy, S. Sen, R. Mahmood, D. Niyogi, M. Lei, S. A. Foster, K. G. Hubbard, E. Douglas, and R. Pielke, 2007:
48 Impacts of the agricultural Green Revolution-induced land use changes on air temperatures in India. *J.*
49 *Geophys. Res. Atmos.*, **112**, 1–13, doi:10.1029/2007JD008834.
- 50 Rubel, F., and M. Kottek, 2010: Observed and projected climate shifts 1901–2100 depicted by world maps of
51 the Köppen-Geiger climate classification. *Meteorol. Zeitschrift*, **19**, 135–141, doi:10.1127/0941-
52 2948/2010/0430.
53 [http://www.schweizerbart.de/papers/metz/detail/19/74932/Observed_and_projected_climate_shifts_19](http://www.schweizerbart.de/papers/metz/detail/19/74932/Observed_and_projected_climate_shifts_1901_2100_de?af=crossref)
54 [01_2100_de?af=crossref](http://www.schweizerbart.de/papers/metz/detail/19/74932/Observed_and_projected_climate_shifts_1901_2100_de?af=crossref) (Accessed April 6, 2019).
- 55 ———, K. Brugger, K. Haslinger, and I. Auer, 2017: The climate of the European Alps: Shift of very high
56 resolution Köppen-Geiger climate zones 1800–2100. *Meteorol. Zeitschrift*, **26**, 115–125,

- 1 doi:10.1127/metz/2016/0816.
2 http://www.schweizerbart.de/papers/metz/detail/26/87237/The_climate_of_the_European_Alps_Shift_of_very_hig?af=crossref (Accessed April 15, 2019).
- 3
4 Rumpel, C., and I. Kögel-Knabner, 2011: Deep soil organic matter—a key but poorly understood component
5 of terrestrial C cycle. *Plant Soil*, **338**, 143–158, doi:10.1007/s11104-010-0391-5.
6 <https://doi.org/10.1007/s11104-010-0391-5>.
- 7 Rumpf, S. B., and Coauthors, 2018: Range dynamics of mountain plants decrease with elevation. *Proc. Natl.*
8 *Acad. Sci.*, **115**, 1848–1853, doi:10.1073/pnas.1713936115.
- 9 Russo, S., and Coauthors, 2014: Magnitude of extreme heat waves in present climate and their projection in
10 a warming world. *J. Geophys. Res. Atmos.*, **119**, 12500–12512, doi:10.1002/2014JD022098.
11 <http://doi.wiley.com/10.1002/2014JD022098> (Accessed October 28, 2018).
- 12 —, A. F. Marchese, J. Sillmann, and G. Imme, 2016: When will unusual heat waves become normal in a
13 warming Africa? *Environ. Res. Lett.*, **11**, 54016, doi:10.1088/1748-9326/11/5/054016.
- 14 Rusticucci, M., 2012: Observed and simulated variability of extreme temperature events over South
15 America. *Atmos. Res.*, **106**, 1–17, doi:10.1016/j.atmosres.2011.11.001.
16 <http://linkinghub.elsevier.com/retrieve/pii/S0169809511003619> (Accessed October 28, 2018).
- 17 —, J. Kyselý, G. Almeida, and O. Lhotka, 2016: Long-term variability of heat waves in Argentina and
18 recurrence probability of the severe 2008 heat wave in Buenos Aires. *Theor. Appl. Climatol.*, **124**, 679–
19 689, doi:10.1007/s00704-015-1445-7. <http://link.springer.com/10.1007/s00704-015-1445-7> (Accessed
20 October 28, 2018).
- 21 Rutting, T., 2017: Nitrogen mineralization, not N₂ fixation, alleviates progressive nitrogen limitation -
22 Comment on “Processes regulating progressive nitrogen limitation under elevated carbon dioxide: a
23 meta-analysis” by Liang et al. (2016). *Biogeosciences*, **14**, 751–754, doi:10.5194/bg-14-751-2017.
- 24 de Sá, S. S., and Coauthors, 2018: Urban influence on the concentration and composition of submicron
25 particulate matter in central Amazonia. *Atmos. Chem. Phys. Discuss.*, doi:10.5194/acp-2018-172.
- 26 —, and Coauthors, 2019: Contributions of biomass-burning, urban, and biogenic emissions to the
27 concentrations and light-absorbing properties of particulate matter in central Amazonia during the dry
28 season. *Atmos. Chem. Phys. Discuss.*, 1–77, doi:10.5194/acp-2018-1309.
- 29 De Sá, S. S., and Coauthors, 2017: Influence of urban pollution on the production of organic particulate
30 matter from isoprene epoxydiols in central Amazonia. *Atmos. Chem. Phys.*, **17**, 6611–6629,
31 doi:10.5194/acp-17-6611-2017.
- 32 Saatchi, S., and Coauthors, 2015: Seeing the forest beyond the trees. *Glob. Ecol. Biogeogr.*,
33 doi:10.1111/geb.12256.
- 34 Saatchi, S. S., and Coauthors, 2011: Benchmark map of forest carbon stocks in tropical regions across three
35 continents. *Proc. Natl. Acad. Sci.*, **108**, 9899–9904, doi:10.1073/pnas.1019576108.
36 <http://www.pnas.org/content/108/24/9899>.
- 37 Sachindra, D. A., A. W. M. M. Ng, S. Muthukumaran, and B. J. C. C. Perera, 2016: Impact of climate
38 change on urban heat island effect and extreme temperatures: a case-study. *Q. J. R. Meteorol. Soc.*, **142**,
39 172–186, doi:10.1002/qj.2642.
- 40 Saeki, T., and P. K. Patra, 2017: Implications of overestimated anthropogenic CO₂ emissions on East Asian
41 and global land CO₂ flux inversion. *Geosci. Lett.*, **4**, 9, doi:10.1186/s40562-017-0074-7.
- 42 Sagan, C., O. B. Toon, and J. B. Pollack, 1979: Anthropogenic Albedo Changes and the Earth’s Climate.
43 *Science (80-)*, **206**, 1363–1368. <http://www.jstor.org/stable/1748990>.
- 44 Sagayama, H., K. Taniguchi, N. Abe, T. H. Arima, M. Soda, M. Matsuura, and K. Hirota, 2008: Correlation
45 between ferroelectric polarization and sense of helical spin order in multiferroic MnWO₄. *Physical*
46 *Review B - Condensed Matter and Materials Physics*, S. Solomon, D. Qin, M. Manning, Z. Chen, M.
47 Marquis, K.B. Averyt, M. Tignor, and H.L. Miller, Eds., Vol. 77 of, Cambridge University Pres,
48 Cambridge, United Kingdom and New York, NY, USA, 129–234
49 <http://en.scientificcommons.org/23467316>.
- 50 Saikawa, E., and Coauthors, 2014: Global and regional emissions estimates for N₂O. *Atmos. Chem. Phys.*,
51 **14**, 4617–4641, doi:10.5194/acp-14-4617-2014.
- 52 Samset, B. H., and Coauthors, 2014a: Modelled black carbon radiative forcing and atmospheric lifetime in
53 AeroCom Phase II constrained by aircraft observations. *Atmos. Chem. Phys.*, **14**, 12465–12477,
54 doi:10.5194/acp-14-12465-2014.
- 55 —, and Coauthors, 2014b: Modelled black carbon radiative forcing and atmospheric lifetime in AeroCom
56 Phase II constrained by aircraft observations. *Atmos. Chem. Phys.*, **14**, 12465–12477, doi:10.5194/acp-

- 1 14-12465-2014.
- 2 Sánchez-Rodríguez, A. R., P. W. Hill, D. R. Chadwick, and D. L. Jones, 2017: Crop residues exacerbate the
3 negative effects of extreme flooding on soil quality. *Biol. Fertil. Soils*, **53**, 751–765,
4 doi:10.1007/s00374-017-1214-0. <http://link.springer.com/10.1007/s00374-017-1214-0> (Accessed May
5 25, 2018).
- 6 Sánchez-Rodríguez, A. R., C. Nie, P. W. Hill, D. R. Chadwick, and D. L. Jones, 2019: Extreme flood events
7 at higher temperatures exacerbate the loss of soil functionality and trace gas emissions in grassland.
8 *Soil Biol. Biochem.*, **130**, 227–236, doi:10.1016/j.soilbio.2018.12.021.
- 9 Sand, M., T. K. Berntsen, K. von Salzen, M. G. Flanner, J. Langner, and D. G. Victor, 2015: Response of
10 Arctic temperature to changes in emissions of short-lived climate forcers. *Nat. Clim. Chang.*, **6**, 286.
11 <https://doi.org/10.1038/nclimate2880>.
- 12 Sanderman, J., T. Hengl, and G. J. Fiske, 2017: Soil carbon debt of 12,000 years of human land use. *Proc.*
13 *Natl. Acad. Sci.*, **114**, 9575–9580, doi:10.1073/pnas.1706103114.
14 <http://www.pnas.org/lookup/doi/10.1073/pnas.1706103114>.
- 15 Sanderson, M. G., C. D. Jones, W. J. Collins, C. E. Johnson, and R. G. Derwent, 2003: Effect of Climate
16 Change on Isoprene Emissions and Surface Ozone Levels. *Geophys. Res. Lett.*, **30**,
17 doi:10.1029/2003GL017642. <https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1029/2003GL017642>.
- 18 Santamouris, M., C. Cartalis, A. Synnefa, and D. Kolokotsa, 2015: On the impact of urban heat island and
19 global warming on the power demand and electricity consumption of buildings - A review. *Energy*
20 *Build.*, **98**, 119–124, doi:10.1016/j.enbuild.2014.09.052.
21 <http://linkinghub.elsevier.com/retrieve/pii/S0378778814007907>.
- 22 Sardans, J., and J. Peñuelas, 2015: Potassium: A neglected nutrient in global change. *Glob. Ecol. Biogeogr.*,
23 **24**, 261–275, doi:10.1111/geb.12259.
- 24 Sardans, J., A. Rivas-Ubach, and J. Peñuelas, 2012: The C:N:P stoichiometry of organisms and ecosystems
25 in a changing world: A review and perspectives. *Perspect. Plant Ecol. Evol. Syst.*, **14**, 33–47,
26 doi:10.1016/j.ppees.2011.08.002.
- 27 Sasaki, N., and Coauthors, 2016: Sustainable Management of Tropical Forests Can Reduce Carbon
28 Emissions and Stabilize Timber Production. *Front. Environ. Sci.*, **4**, 50, doi:10.3389/fenvs.2016.00050.
29 <http://journal.frontiersin.org/Article/10.3389/fenvs.2016.00050/abstract>.
- 30 Sathre, R., and J. O'Connor, 2010: Meta-analysis of greenhouse gas displacement factors of wood product
31 substitution. *Environ. Sci. Policy*, **13**, 104–114, doi:10.1016/j.envsci.2009.12.005.
32 <http://www.sciencedirect.com/science/article/pii/S1462901109001804>.
- 33 Satria, 2017: Environmental Research Letters Substantial N₂O emissions from peat decomposition and N
34 fertilization in an oil palm plantation exacerbated by hotspots. *Environ. Res. Lett.*, **12**,
35 doi:<https://doi.org/10.1088/1748-9326/aa80f1>.
- 36 Saturno, J., and Coauthors, 2018: Black and brown carbon over central Amazonia: Long-term aerosol
37 measurements at the ATTO site. *Atmos. Chem. Phys.*, **18**, 12817–12843, doi:10.5194/acp-18-12817-
38 2018.
- 39 Savage, J., and M. Vellend, 2015: Elevational shifts, biotic homogenization and time lags in vegetation
40 change during 40 years of climate warming. *Ecography (Cop.)*, **38**, 546–555, doi:10.1111/ecog.01131.
41 <http://doi.wiley.com/10.1111/ecog.01131> (Accessed October 31, 2018).
- 42 Sayer, E. J., M. S. Heard, H. K. Grant, T. R. Marthews, and E. V. J. Tanner, 2011: Soil carbon release
43 enhanced by increased tropical forest litterfall. *Nat. Clim. Chang.*, **1**, 304–307,
44 doi:10.1038/nclimate1190. <http://dx.doi.org/10.1038/nclimate1190>.
- 45 Scalenghe, R., and F. A. Marsan, 2009: The anthropogenic sealing of soils in urban areas. *Landsc. Urban*
46 *Plan.*, **90**, 1–10, doi:<https://doi.org/10.1016/j.landurbplan.2008.10.011>.
47 <http://www.sciencedirect.com/science/article/pii/S0169204608001710>.
- 48 Schaefer, H., and Coauthors, 2016: A 21st-century shift from fossil-fuel to biogenic methane emissions
49 indicated by 13CH₄. *Science (80-.)*, **352**, 80–84, doi:10.1126/science.aad2705.
- 50 Schafer, J. S., T. F. Eck, B. N. Holben, P. Artaxo, M. A. Yamasoe, and A. S. Procopio, 2002: Observed
51 reductions of total solar irradiance by biomass-burning aerosols in the Brazilian Amazon and Zambian
52 Savanna. *Geophys. Res. Lett.*, **29**, 4-1-4–4, doi:10.1029/2001GL014309.
53 <http://doi.wiley.com/10.1029/2001GL014309> (Accessed May 6, 2011).
- 54 Schär, C., 2016: Climate extremes: The worst heat waves to come. *Nat. Clim. Chang.*, **6**, 128–129,
55 doi:10.1038/nclimate2864. <http://www.nature.com/articles/nclimate2864> (Accessed October 28, 2018).
- 56 Scherer, M., and N. S. Diffenbaugh, 2014: Transient twenty-first century changes in daily-scale temperature

- 1 extremes in the United States. *Clim. Dyn.*, **42**, 1383–1404, doi:10.1007/s00382-013-1829-2.
2 <http://link.springer.com/10.1007/s00382-013-1829-2> (Accessed October 28, 2018).
- 3 Schimel, D., B. B. Stephens, and J. B. Fisher, 2015: Effect of increasing CO₂ on the terrestrial carbon cycle.
4 *Proc. Natl. Acad. Sci.*, **112**, 436–441, doi:10.1073/pnas.1407302112.
5 <http://www.pnas.org/lookup/doi/10.1073/pnas.1407302112>.
- 6 Schindlbacher, A., J. Schnecker, M. Takriti, W. Borcken, and W. Wanek, 2015: Microbial physiology and soil
7 CO₂efflux after 9 years of soil warming in a temperate forest - no indications for thermal adaptations.
8 *Glob. Chang. Biol.*, **21**, 4265–4277, doi:10.1111/gcb.12996.
- 9 Schlaepfer, D. R., and Coauthors, 2017: Climate change reduces extent of temperate drylands and intensifies
10 drought in deep soils. *Nat. Commun.*, **8**, 14196, doi:10.1038/ncomms14196.
11 <http://www.nature.com/doi/10.1038/ncomms14196> (Accessed October 15, 2018).
- 12 Schlenker, W., and D. B. Lobell, 2010: Robust negative impacts of climate change on African agriculture.
13 *Environ. Res. Lett.*, **5**, 14010, doi:10.1088/1748-9326/5/1/014010.
- 14 Schlesinger, W. H., M. C. Dietze, R. B. Jackson, R. P. Phillips, C. C. Rhoades, L. E. Rustad, and J. M. Vose,
15 2016: Forest biogeochemistry in response to drought. *Glob. Chang. Biol.*, **22**, 2318–2328,
16 doi:10.1111/gcb.13105. <http://doi.wiley.com/10.1111/gcb.13105> (Accessed May 7, 2018).
- 17 Schleussner, C.-F., P. Pfliederer, and E. M. Fischer, 2017: In the observational record half a degree matters.
18 *Nat. Clim. Chang.*, **7**, 460–462, doi:10.1038/nclimate3320.
19 <http://www.nature.com/articles/nclimate3320> (Accessed March 24, 2019).
- 20 Schlünzen, K. H., P. Hoffmann, G. Rosenhagen, and W. Riecke, 2010: Long-term changes and regional
21 differences in temperature and precipitation in the metropolitan area of Hamburg. *Int. J. Climatol.*, **30**,
22 1121–1136, doi:10.1002/joc.1968.
- 23 Schmale, J., and Coauthors, 2018: Long-term cloud condensation nuclei number concentration, particle
24 number size distribution and chemical composition measurements at regionally representative
25 observatories. *Atmos. Chem. Phys.*, **18**, 2853–2881, doi:10.5194/acp-18-2853-2018.
- 26 Schmidt, M., 2011: Persistence of soil organic matter as an ecosystem property. *Nature*, **478**, 49.
27 <http://www.nature.com/nature/journal/v478/n7367/pdf/nature10386.pdf>.
- 28 Schueler, V., S. Fuss, J. C. Steckel, U. Weddige, and T. Beringer, 2016: Productivity ranges of sustainable
29 biomass potentials from non-agricultural land. *Environ. Res. Lett.*, doi:10.1088/1748-
30 9326/11/7/074026.
- 31 Schulte-Uebbing, L., and W. de Vries, 2018: Global-scale impacts of nitrogen deposition on tree carbon
32 sequestration in tropical, temperate, and boreal forests: A meta-analysis. *Glob. Chang. Biol.*, **24**, e416–
33 e431, doi:10.1111/gcb.13862.
- 34 Schultz, N. M., P. J. Lawrence, and X. Lee, 2017: Global satellite data highlights the diurnal asymmetry of
35 the surface temperature response to deforestation. *J. Geophys. Res. Biogeosciences*, **122**, 903–917,
36 doi:10.1002/2016JG003653.
- 37 Schulz, C., and Coauthors, 2018: Aircraft-based observations of isoprene epoxydiol-derived secondary
38 organic aerosol (IEPOX-SOA) in the tropical upper troposphere over the Amazon region. *Atmos.*
39 *Chem. Phys. Discuss.*, **3**, 1–32, doi:10.5194/acp-2018-232.
- 40 Schuur, E. A. G., and Coauthors, 2015a: Climate change and the permafrost carbon feedback. *Nature*, **520**,
41 171–179, doi:10.1038/nature14338. <http://dx.doi.org/10.1038/nature14338> (Accessed October 25,
42 2018).
- 43 Schuur, E. A. G., and Coauthors, 2015b: Climate change and the permafrost carbon feedback. *Nature*, **520**,
44 171–179, doi:10.1038/nature14338. <http://dx.doi.org/10.1038/nature14338>.
- 45 Scott, A. ., 2000: The Pre-Quaternary history of fire. *Palaeogeogr. Palaeoclimatol. Palaeoecol.*, **164**, 281–
46 329, doi:10.1016/S0031-0182(00)00192-9.
47 <http://linkinghub.elsevier.com/retrieve/pii/S0031018200001929> (Accessed May 7, 2018).
- 48 Scott, C. E., and Coauthors, 2015: Impact of gas-to-particle partitioning approaches on the simulated
49 radiative effects of biogenic secondary organic aerosol. *Atmos. Chem. Phys.*, **15**, 12989–13001,
50 doi:10.5194/acp-15-12989-2015.
- 51 ———, and Coauthors, 2017: Impact on short-lived climate forcers (SLCFs) from a realistic land-use
52 change scenario via changes in biogenic emissions. *Faraday Discuss.*, **200**, 101–120,
53 doi:10.1039/c7fd00028f.
- 54 Scott, C. E., and Coauthors, 2018a: Impact on short-lived climate forcers increases projected warming due to
55 deforestation. *Nat. Commun.*, **9**, 157, doi:10.1038/s41467-017-02412-4.
- 56 Scott, C. E., S. R. Arnold, S. A. Monks, A. Asmi, P. Paasonen, and D. V. Spracklen, 2018b: Substantial

- 1 large-scale feedbacks between natural aerosols and climate. *Nat. Geosci.*, **11**, 44–48,
2 doi:10.1038/s41561-017-0020-5.
- 3 Searle, S., and C. Malins, 2015: A reassessment of global bioenergy potential in 2050. *GCB Bioenergy*,
4 doi:10.1111/gcbb.12141.
- 5 Sebastien Fontaine Pierre Barre, Nadia Bdioui, Bruno Mary, Cornelia Rumpel, S. B., 2007: Stability of
6 organic carbon in deep soil layers. *Nature*, **450**, 277–281. <http://dx.doi.org/10.1038/nature06275>.
- 7 Seddon, A. W. R., M. Macias-Fauria, P. R. Long, D. Benz, and K. J. Willis, 2016: Sensitivity of global
8 terrestrial ecosystems to climate variability. *Nature*, **531**, 229–232, doi:10.1038/nature16986.
- 9 Seidl, R., and Coauthors, 2017: Forest disturbances under climate change. *Nat. Clim. Chang.*, **7**, 395–402,
10 doi:10.1038/nclimate3303.
- 11 Sejas, S. A., O. S. Albert, M. Cai, and Y. Deng, 2014: Feedback attribution of the land-sea warming contrast
12 in a global warming simulation of the NCAR CCSM4. *Environ. Res. Lett.*, **9**, doi:10.1088/1748-
13 9326/9/12/124005.
- 14 Sellers, P. J., and Coauthors, 1996: A revised land surface parameterization (SiB2) for atmospheric GCMs.
15 Part I: Model formulation. *J. Clim.*, **9**, 676–705, doi:10.1175/1520-
16 0442(1996)009<0676:ARLSPF>2.0.CO;2. <https://doi.org/10.1175/1520->
17 [0442\(1996\)009%3C0676:ARLSPF%3E2.0.CO](https://doi.org/10.1175/1520-0442(1996)009%3C0676:ARLSPF%3E2.0.CO).
- 18 Semazzi, F. H. M., and Y. Song, 2001: A GCM study of climate change induced by deforestation in Africa.
19 *Clim. Res.*, **17**, 169–182, doi:10.3354/cr017169.
- 20 Sena, E. T., A. McComiskey, and G. Feingold, 2016: A long-term study of aerosol-cloud interactions and
21 their radiative effect at the Southern Great Plains using ground-based measurements. *Atmos. Chem.*
22 *Phys.*, **16**, 11301–11318, doi:10.5194/acp-16-11301-2016.
- 23 Seneviratne, S. I., D. Lüthi, M. Litschi, and C. Schär, 2006: Land-atmosphere coupling and climate change
24 in Europe. *Nature*, **443**, 205–209, doi:10.1038/nature05095.
25 <http://www.nature.com/doi/finder/10.1038/nature05095> (Accessed November 1, 2018).
- 26 Seneviratne, S. I., and Coauthors, 2012a: Changes in climate extremes and their impacts on the natural
27 physical environment. *Managing the Risks of Extreme Events and Disasters to Advance Climate*
28 *Change Adaptation. A Special Report of Working Groups I and II of the Intergovernmental Panel on*
29 *Climate Change (IPCC)*, C.B. Field et al., Eds., Cambridge University Press, Cambridge, United
30 Kingdom and New York, NY, USA, 109–230 <http://ipcc-wg2.gov/SREX/images/uploads/SREX->
31 [Chap3_FINAL.pdf](http://ipcc-wg2.gov/SREX/images/uploads/SREX-Chap3_FINAL.pdf).
- 32 Seneviratne, S. I., and Coauthors, 2013: Impact of soil moisture-climate feedbacks on CMIP5 projections:
33 First results from the GLACE-CMIP5 experiment. *Geophys. Res. Lett.*, **40**, 5212–5217,
34 doi:10.1002/grl.50956.
- 35 ———, M. G. Donat, A. J. Pitman, R. Knutti, and R. L. Wilby, 2016: Allowable CO₂ emissions based on
36 regional and impact-related climate targets. *Nature*, **529**, 477–483, doi:10.1038/nature16542.
- 37 ———, and Coauthors, 2018: Land radiative management as contributor to regional-scale climate adaptation
38 and mitigation. *Nat. Geosci.*, **11**, 88–96, doi:10.1038/s41561-017-0057-5.
- 39 Seneviratne, S. I. S. I., and Coauthors, 2012b: Changes in climate extremes and their impacts on the natural
40 physical environment. *Managing the Risks of Extreme Events and Disasters to Advance Climate*
41 *Change Adaptation. A Special Report of Working Groups I and II of the Intergovernmental Panel on*
42 *Climate Change (IPCC)*, C.B. Field et al., Eds., Cambridge University Press, Cambridge, United
43 Kingdom and New York, NY, USA, 109–230 <http://ipcc-wg2.gov/SREX/images/uploads/SREX->
44 [Chap3_FINAL.pdf](http://ipcc-wg2.gov/SREX/images/uploads/SREX-Chap3_FINAL.pdf).
- 45 Seto, K. C., M. Fragkias, B. Güneralp, and M. K. Reilly, 2011: A meta-analysis of global urban land
46 expansion. *PLoS One*, **6**, e23777, doi:10.1371/journal.pone.0023777.
- 47 Settele, J., and Coauthors, 2015a: Terrestrial and Inland water systems. *Climate Change 2014 Impacts,*
48 *Adaptation and Vulnerability: Part A: Global and Sectoral Aspects*, C.B. Field et al., Eds., Cambridge
49 University Press, Cambridge, United Kingdom and New York, NY, USA, 271–360
50 <http://ebooks.cambridge.org/ref/id/CBO9781107415379A025>.
- 51 ———, and Coauthors, 2015b: Terrestrial and Inland water systems. *Climate Change 2014 Impacts,*
52 *Adaptation and Vulnerability: Part A: Global and Sectoral Aspects*, C.B. Field et al., Eds., Cambridge
53 University Press, Cambridge, United Kingdom and New York, NY, USA, United Kingdom and New
54 York, NY, USA, 271–360 <http://ebooks.cambridge.org/ref/id/CBO9781107415379A025>.
- 55 Shade, C., and P. Kremer, 2019: Predicting Land Use Changes in Philadelphia Following Green
56 Infrastructure Policies. *Land*, **8**, doi:10.3390/land8020028. <http://www.mdpi.com/2073-445X/8/2/28>.

- 1 Shao, P., and Coauthors, 2013: Soil microbial respiration from observations and Earth System Models.
2 *Environ. Res. Lett.*, **8**, doi:10.1088/1748-9326/8/3/034034.
- 3 Sharma, S. B., S. Jain, P. Khirwadkar, and S. Kulkarni, 2013: The effects of air pollution on the environment
4 and human health. *Indian J. Res. Pharm. Biotechnol.*, **1**, 2320–3471.
- 5 Shcherbak, I., N. Millar, and G. P. Robertson, 2014: Global metaanalysis of the nonlinear response of soil
6 nitrous oxide (N₂O) emissions to fertilizer nitrogen. *Proc. Natl. Acad. Sci.*, **111**, 9199–9204,
7 doi:10.1073/pnas.1322434111. <http://www.pnas.org/cgi/doi/10.1073/pnas.1322434111>.
- 8 Sheffield, J., E. F. Wood, and M. L. Roderick, 2012: Little change in global drought over the past 60 years.
9 *Nature*, **491**, 435–438, doi:10.1038/nature11575. <http://www.nature.com/articles/nature11575>
10 (Accessed May 7, 2018).
- 11 Shevliakova, E., R. J. Stouffer, S. Malyshev, J. P. Krasting, G. C. Hurtt, and S. W. Pacala, 2013: Historical
12 warming reduced due to enhanced land carbon uptake. *Proc. Natl. Acad. Sci.*, **110**, 16730–16735,
13 doi:10.1073/pnas.1314047110.
- 14 Shi, W., F. Tao, and J. Liu, 2013: Regional temperature change over the Huang-Huai-Hai Plain of China: the
15 roles of irrigation versus urbanization. *Int. J. Climatol.*, **34**, 1181–1195, doi:10.1002/joc.3755.
16 <https://doi.org/10.1002/joc.3755>.
- 17 Shindell, D., and Coauthors, 2013: 2013: Anthropogenic and Natural Radiative Forcing. *Climate Change*
18 *2013: The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of*
19 *the Intergovernmental Panel on Climate Change*, T.F. Stocker et al., Eds., Vol. 9781107057 of,
20 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA, 659–740
21 <http://www.climatechange2013.org>.
- 22 Shindell, D., and Coauthors, 2017: A climate policy pathway for near- and long-term benefits. *Science (80-*
23 *),* **356**, 493–494, doi:10.1126/science.aak9521.
- 24 Shindell, D., G. Faluvegi, K. Seltzer, and C. Shindell, 2018: Quantified, localized health benefits of
25 accelerated carbon dioxide emissions reductions. *Nat. Clim. Chang.*, **8**, 291–295, doi:10.1038/s41558-
26 018-0108-y. <https://doi.org/10.1038/s41558-018-0108-y>.
- 27 Siebert, S., V. Henrich, K. Frenken, and J. Burke, 2013: Global Map of Irrigation Areas version 5. *Univ.*
28 *Frankfurt (Main), Ger. FAO, Rome, Italy.*, 1–2, doi:10.13140/2.1.2660.6728.
- 29 ———, H. Webber, G. Zhao, and F. Ewert, 2017: Heat stress is overestimated in climate impact studies for
30 irrigated agriculture. *Environ. Res. Lett.*, **12**, 54023, doi:10.1088/1748-9326/aa702f.
31 <http://stacks.iop.org/1748-9326/12/i=5/a=054023?key=crossref.a46b95ee7233b45a96fada3a99ce956a>
32 (Accessed May 21, 2018).
- 33 Silva, L. C. R., and M. Anand, 2013: Probing for the influence of atmospheric CO₂ and climate change on
34 forest ecosystems across biomes. *Glob. Ecol. Biogeogr.*, **22**, 83–92, doi:10.1111/j.1466-
35 8238.2012.00783.x.
- 36 Simmons, C. T., and H. D. Matthews, 2016: Assessing the implications of human land-use change for the
37 transient climate response to cumulative carbon emissions. *Environ. Res. Lett.*, **11**, doi:10.1088/1748-
38 9326/11/3/035001.
- 39 Singarayer, J. S., and T. Davies-Barnard, 2012: Regional climate change mitigation with crops: Context and
40 assessment. *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, **370**, 4301–4316,
41 doi:10.1098/rsta.2012.0010.
- 42 Singarayer, J. S., A. Ridgwell, and P. Irvine, 2009: Assessing the benefits of crop albedo bio-geoengineering.
43 *Environ. Res. Lett.*, **4**, doi:10.1088/1748-9326/4/4/045110.
- 44 Singh, B. K., R. D. Bardgett, P. Smith, and D. S. Reay, 2010: Microorganisms and climate change:
45 Terrestrial feedbacks and mitigation options. *Nat. Rev. Microbiol.*, **8**, 779–790,
46 doi:10.1038/nrmicro2439. <http://dx.doi.org/10.1038/nrmicro2439>.
- 47 Singh, D., M. G. Flanner, and J. Perket, 2015: The global land shortwave cryosphere radiative effect during
48 the MODIS era. *Cryosphere*, **9**, 2057–2070, doi:10.5194/tc-9-2057-2015.
- 49 Sippel, S., and F. E. L. Otto, 2014: Beyond climatological extremes - assessing how the odds of
50 hydrometeorological extreme events in South-East Europe change in a warming climate. *Clim. Change*,
51 **125**, 381–398, doi:10.1007/s10584-014-1153-9. <http://link.springer.com/10.1007/s10584-014-1153-9>
52 (Accessed May 25, 2018).
- 53 ———, J. Zscheischler, M. D. Mahecha, R. Orth, M. Reichstein, M. Vogel, and S. I. Seneviratne, 2017:
54 Refining multi-model projections of temperature extremes by evaluation against
55 land&ndash;atmosphere coupling diagnostics. *Earth Syst. Dyn.*, **8**, 387–403, doi:10.5194/esd-
56 8-387-2017.

- 1 Sirin, A.A., Suvorov, G.G., Chistotin, M.V., Glagolev, M. V., 2012: About Meaning of Methane Emission
2 From Drainage Channels. *Doosigik*, **2**, 1–10.
- 3 Sirin, A., T. Minayeva, A. Vozbrannaya, and S. Bartalev, 2011: How to avoid peat fires? *Sci. Russ.*,
4 Sisti, C. P. J., H. P. Dos Santos, R. Kohhann, B. J. R. Alves, S. Urquiaga, and R. M. Boddey, 2004: Change
5 in carbon and nitrogen stocks in soil under 13 years of conventional or zero tillage in southern Brazil.
6 *Soil Tillage Res.*, **76**, 39–58, doi:10.1016/j.still.2003.08.007.
- 7 Sitch, S., V. Brovkin, W. von Bloh, D. van Vuuren, B. Eickhout, and A. Ganopolski, 2005: Impacts of future
8 land cover changes on atmospheric CO₂ and climate. *Global Biogeochem. Cycles*, **19**, 1–15,
9 doi:10.1029/2004GB002311.
- 10 Sitch, S., and Coauthors, 2015a: Recent trends and drivers of regional sources and sinks of carbon dioxide.
11 *Biogeosciences*, **12**, 653–679, doi:10.5194/bg-12-653-2015.
12 <https://www.biogeosciences.net/12/653/2015/>.
- 13 Sitch, S., and Coauthors, 2015b: Recent trends and drivers of regional sources and sinks of carbon dioxide.
14 *Biogeosciences*, **12**, 653–679, doi:10.5194/bg-12-653-2015.
15 <https://www.biogeosciences.net/12/653/2015/>.
- 16 Six, J., R. T. Conant, E. A. Paul, and K. Paustian, 2002: Stabilization mechanisms of soil organic matter:
17 Implications for C-saturation of soils. *Plant Soil*, **241**, 155–176, doi:10.1023/A:1016125726789.
- 18 Six, J., S. M. Ogle, F. J. Breidt, R. T. Conant, A. R. Mosiers, and K. Paustian, 2004: The potential to mitigate
19 global warming with no-tillage management is only realised when practised in the long term. *Glob.*
20 *Chang. Biol.*, **10**, 155–160, doi:10.1111/j.1529-8817.2003.00730.x.
- 21 van der Sleen, P., and Coauthors, 2014: No growth stimulation of tropical trees by 150 years of CO₂
22 fertilization but water-use efficiency increased. *Nat. Geosci.*, **8**, 24–28, doi:10.1038/ngeo2313.
- 23 Slot, M., and K. Kitajima, 2015: General patterns of acclimation of leaf respiration to elevated temperatures
24 across biomes and plant types. *Oecologia*, **177**, 885–900, doi:10.1007/s00442-014-3159-4.
- 25 —, and K. Winter, 2017: Photosynthetic acclimation to warming in tropical forest tree Seedlings. *J. Exp.*
26 *Bot.*, **68**, 2275–2284, doi:10.1093/jxb/erx071.
- 27 —, C. Rey-Sánchez, S. Gerber, J. W. Lichstein, K. Winter, and K. Kitajima, 2014: Thermal acclimation of
28 leaf respiration of tropical trees and lianas: Response to experimental canopy warming, and
29 consequences for tropical forest carbon balance. *Glob. Chang. Biol.*, **20**, 2915–2926,
30 doi:10.1111/gcb.12563.
- 31 Smith, M. C., J. S. Singarayer, P. J. Valdes, J. O. Kaplan, and N. P. Branch, 2016a: The biogeophysical
32 climatic impacts of anthropogenic land use change during the Holocene. *Clim. Past*, **12**, 923–941,
33 doi:10.5194/cp-12-923-2016.
- 34 Smith, N. G., and J. S. Dukes, 2013: Plant respiration and photosynthesis in global-scale models:
35 Incorporating acclimation to temperature and CO₂. *Glob. Chang. Biol.*, **19**, 45–63, doi:10.1111/j.1365-
36 2486.2012.02797.x.
- 37 —, S. L. Malyshev, E. Shevliakova, J. Kattge, and J. S. Dukes, 2016b: Foliar temperature acclimation
38 reduces simulated carbon sensitivity to climate. *Nat. Clim. Chang.*, **6**, 407–411,
39 doi:10.1038/nclimate2878.
- 40 Smith, P., 2016: Soil carbon sequestration and biochar as negative emission technologies. *Glob. Chang.*
41 *Biol.*, **22**, 1315–1324, doi:10.1111/gcb.13178. <https://doi.org/10.1111/gcb.13178>.
- 42 —, and Coauthors, 2008: Greenhouse gas mitigation in agriculture. *Philos. Trans. R. Soc. B Biol. Sci.*,
43 **363**, 789–813, doi:10.1098/rstb.2007.2184.
44 <http://rstb.royalsocietypublishing.org/cgi/doi/10.1098/rstb.2007.2184>.
- 45 —, and Coauthors, 2013a: How much land-based greenhouse gas mitigation can be achieved without
46 compromising food security and environmental goals? *Glob. Chang. Biol.*, **19**, 2285–2302,
47 doi:10.1111/gcb.12160. <https://doi.org/10.1111/gcb.12160>.
- 48 —, and Coauthors, 2015: Global change pressures on soils from land use and management. *Glob. Chang.*
49 *Biol.*, **22**, 1008–1028, doi:10.1111/gcb.13068. <http://doi.wiley.com/10.1111/gcb.13068> (Accessed May
50 25, 2018).
- 51 —, and Coauthors, 2016c: Biophysical and economic limits to negative CO₂ emissions. *Nat. Clim. Chang.*,
52 **6**, 42–50, doi:10.1038/nclimate2870. <http://dx.doi.org/10.1038/nclimate2870>.
- 53 —, and Coauthors, 2016d: Biophysical and economic limits to negative CO₂ emissions. *Nat. Clim.*
54 *Chang.*, **6**, 42–50.
- 55 Smith, S. V, W. H. Renwick, R. W. Buddemeier, and C. J. Crossland, 2001: Budgets of soil erosion and
56 deposition for sediments and sedimentary organic carbon across the conterminous United States.

- 1 *Global Biogeochem. Cycles*, **15**, 697–707, doi:10.1029/2000GB001341.
2 <https://doi.org/10.1029/2000GB001341>.
- 3 Smith, T. T., B. F. Zaitchik, and J. M. Gohlke, 2013b: Heat waves in the United States: Definitions, patterns
4 and trends. *Clim. Change*, **118**, 811–825, doi:10.1007/s10584-012-0659-2.
5 <http://link.springer.com/10.1007/s10584-012-0659-2> (Accessed May 16, 2018).
- 6 Smyth, C., W. A. Kurz, G. Rampley, T. C. Lemprière, and O. Schwab, 2017a: Climate change mitigation
7 potential of local use of harvest residues for bioenergy in Canada. *GCB Bioenergy*, **9**, 817–832,
8 doi:10.1111/gcbb.12387.
- 9 ———, G. Rampley, T. C. Lemprière, O. Schwab, and W. A. Kurz, 2017b: Estimating product and energy
10 substitution benefits in national-scale mitigation analyses for Canada. *GCB Bioenergy*, **9**, 1071–1084,
11 doi:10.1111/gcbb.12389. <https://doi.org/10.1111/gcbb.12389>.
- 12 Snider, D. M., J. J. Venkiteswaran, S. L. Schiff, and J. Spoelstra, 2015: From the Ground Up: Global Nitrous
13 Oxide Sources are Constrained by Stable Isotope Values. *PLoS One*, **10**, e0118954,
14 doi:10.1371/journal.pone.0118954.
- 15 Snyder, P. K., 2010: The influence of tropical deforestation on the Northern Hemisphere climate by
16 atmospheric teleconnections. *Earth Interact.*, doi:10.1175/2010EI280.1.
- 17 ———, C. Delire, and J. a. Foley, 2004a: Evaluating the influence of different vegetation biomes on the global
18 climate. *Clim. Dyn.*, **23**, 279–302, doi:10.1007/s00382-004-0430-0.
- 19 Snyder, P. K., C. Delire, and J. A. Foley, 2004b: Evaluating the influence of different vegetation biomes on
20 the global climate. *Clim. Dyn.*, **23**, 279–302, doi:10.1007/s00382-004-0430-0.
- 21 Soimakallio, S., L. Saikku, L. Valsta, and K. Pingoud, 2016: Climate Change Mitigation Challenge for
22 Wood Utilization-The Case of Finland. *Environ. Sci. Technol.*, doi:10.1021/acs.est.6b00122.
- 23 Soja, A. J., and Coauthors, 2007: Climate-induced boreal forest change: Predictions versus current
24 observations. *Glob. Planet. Change*, **56**, 274–296, doi:10.1016/j.gloplacha.2006.07.028.
25 <https://www.sciencedirect.com/science/article/pii/S0921818106001883> (Accessed April 2, 2019).
- 26 Sommer, R., and D. Bossio, 2014: Dynamics and climate change mitigation potential of soil organic carbon
27 sequestration. *J. Environ. Manage.*, doi:10.1016/j.jenvman.2014.05.017.
- 28 Son, K. H., and D. H. Bae, 2015: Drought analysis according to shifting of climate zones to arid climate zone
29 over Asia monsoon region. *J. Hydrol.*, **529**, 1021–1029, doi:10.1016/j.jhydrol.2015.09.010.
30 <https://www.sciencedirect.com/science/article/pii/S0022169415006939> (Accessed April 15, 2019).
- 31 Sonntag, S., J. Pongratz, C. H. Reick, and H. Schmidt, 2016: Reforestation in a high-CO₂ world—Higher
32 mitigation potential than expected, lower adaptation potential than hoped for. *Geophys. Res. Lett.*, **43**,
33 6546–6553, doi:10.1002/2016GL068824.
- 34 Sonntag, S., and Coauthors, 2018: Quantifying and Comparing Effects of Climate Engineering Methods on
35 the Earth System. *Earth's Futur.*, doi:10.1002/ef2.285.
- 36 Spahni, R., F. Joos, B. D. Stocker, M. Steinacher, and Z. C. Yu, 2013: Transient simulations of the carbon
37 and nitrogen dynamics in northern peatlands: From the Last Glacial Maximum to the 21st century.
38 *Clim. Past*, doi:10.5194/cp-9-1287-2013.
- 39 Spencer, T., and Coauthors, 2016: Global coastal wetland change under sea-level rise and related stresses:
40 The DIVA Wetland Change Model. *Glob. Planet. Change*, **139**, 15–30,
41 doi:10.1016/j.gloplacha.2015.12.018.
42 <http://www.sciencedirect.com/science/article/pii/S0921818115301879>.
- 43 Sperry, J. S., and D. M. Love, 2015: What plant hydraulics can tell us about responses to climate-change
44 droughts. *New Phytol.*, **207**, 14–27, doi:10.1111/nph.13354.
- 45 Spinoni, J., G. Naumann, J. Vogt, and P. Barbosa, 2015a: European drought climatologies and trends based
46 on a multi-indicator approach. *Glob. Planet. Change*, **127**, 50–57,
47 doi:10.1016/J.GLOPLACHA.2015.01.012.
48 <https://www.sciencedirect.com/science/article/pii/S0921818115000284?via%3Dihub> (Accessed
49 October 29, 2018).
- 50 ———, J. Vogt, G. Naumann, H. Carrao, and P. Barbosa, 2015b: Towards identifying areas at climatological
51 risk of desertification using the Köppen-Geiger classification and FAO aridity index. *Int. J. Climatol.*,
52 **35**, 2210–2222, doi:10.1002/joc.4124. <http://doi.wiley.com/10.1002/joc.4124> (Accessed April 15,
53 2019).
- 54 ———, and Coauthors, 2019: A new global database of meteorological drought events from 1951 to 2016. *J.*
55 *Hydrol. Reg. Stud.*, **22**, 100593, doi:10.1016/j.ejrh.2019.100593.
- 56 Sporre, M. K., S. M. Blichner, I. H. H. Karset, R. Makkonen, and T. K. Berntsen, 2018: BVOC-aerosol-

- 1 climate feedbacks investigated using NorESM. *Atmos. Chem. Phys. Discuss.*, **2018**, 1–34,
2 doi:10.5194/acp-2018-930.
- 3 Spracklen, D. V., J. C. A. Baker, L. Garcia-Carreras, and J. H. Marsham, 2018: The Effects of Tropical
4 Vegetation on Rainfall. *Annu. Rev. Environ. Resour.*, **43**, 193–218, doi:10.1146/annurev-environ-
5 102017-030136.
- 6 Spracklen, D. V., and L. Garcia-Carreras, 2015: The impact of Amazonian deforestation on Amazon basin
7 rainfall. *Geophys. Res. Lett.*, **42**, 9546–9552, doi:10.1002/2015GL066063.
8 <http://doi.wiley.com/10.1002/2015GL066063> (Accessed December 20, 2017).
- 9 ———, S. R. Arnold, and C. M. Taylor, 2012: Observations of increased tropical rainfall preceded by air
10 passage over forests. *Nature*, **489**, 282–285, doi:10.1038/nature11390.
- 11 Springmann, M., H. C. J. Godfray, M. Rayner, and P. Scarborough, 2016: Analysis and valuation of the
12 health and climate change cobenefits of dietary change. *Proc. Natl. Acad. Sci.*, **113**, 4146–4151,
13 doi:10.1073/pnas.1523119113. <http://www.pnas.org/lookup/doi/10.1073/pnas.1523119113>.
- 14 Staal, A., O. A. Tuinenburg, J. H. C. Bosmans, M. Holmgren, E. H. Van Nes, M. Scheffer, D. C. Zemp, and
15 S. C. Dekker, 2018: Forest-rainfall cascades buffer against drought across the Amazon.
16 doi:10.1038/s41558-018-0177-y.
- 17 Stallard, R. F., 1998: Terrestrial sedimentation and the carbon cycle: coupling weathering and erosion to
18 carbon burial. *Global Biogeochem. Cycles*, **12**, 231–257.
- 19 Stanelle, T., I. Bey, T. Raddatz, C. Reick, and I. Tegen, 2014: Anthropogenically induced changes in
20 twentieth century mineral dust burden and the associated impact on radiative forcing. *J. Geophys. Res.*,
21 **119**, 13,526–13,546, doi:10.1002/2014JD022062.
- 22 Stanne, T. M., L. L. E. Sjögren, S. Koussevitzky, and A. K. Clarke, 2009: Identification of new protein
23 substrates for the chloroplast ATP-dependent Clp protease supports its constitutive role in *Arabidopsis*.
24 *Biochem. J.*, **417**, 257–269, doi:10.1042/BJ20081146.
25 <http://biochemj.org/lookup/doi/10.1042/BJ20081146>.
- 26 Staples, M. D., R. Malina, and S. R. H. Barrett, 2017: The limits of bioenergy for mitigating global life-cycle
27 greenhouse gas emissions from fossil fuels. *Nat. Energy*, **2**, 16202, doi:10.1038/nenergy.2016.202
28 <http://www.nature.com/articles/nenergy2016202#supplementary-information>.
29 <http://dx.doi.org/10.1038/nenergy.2016.202>.
- 30 Steinbauer, M. J., and Coauthors, 2018: Accelerated increase in plant species richness on mountain summits
31 is linked to warming. *Nature*, **556**, 231–234, doi:10.1038/s41586-018-0005-6.
32 <http://www.nature.com/articles/s41586-018-0005-6> (Accessed August 30, 2018).
- 33 Stevanović, M., and Coauthors, 2017: Mitigation Strategies for Greenhouse Gas Emissions from Agriculture
34 and Land-Use Change: Consequences for Food Prices. *Environ. Sci. Technol.*, **51**, 365–374,
35 doi:10.1021/acs.est.6b04291. <http://pubs.acs.org/doi/10.1021/acs.est.6b04291> (Accessed May 25,
36 2018).
- 37 Stevens-Rumann, C. S., K. B. Kemp, P. E. Higuera, B. J. Harvey, M. T. Rother, D. C. Donato, P. Morgan,
38 and T. T. Veblen, 2018: Evidence for declining forest resilience to wildfires under climate change.
39 *Ecol. Lett.*, **21**, 243–252, doi:10.1111/ele.12889. <http://doi.wiley.com/10.1111/ele.12889> (Accessed
40 October 30, 2018).
- 41 Stocker, B. D., F. Feissli, K. M. Strassmann, R. Spahni, and F. Joos, 2014: Past and future carbon fluxes
42 from land use change, shifting cultivation and wood harvest. *Tellus, Ser. B Chem. Phys. Meteorol.*,
43 doi:10.3402/tellusb.v66.23188.
- 44 Stott, P. A., and Coauthors, 2016: Attribution of extreme weather and climate-related events. *Wiley*
45 *Interdiscip. Rev. Clim. Chang.*, **7**, 23–41, doi:10.1002/wcc.380. <http://doi.wiley.com/10.1002/wcc.380>
46 (Accessed October 28, 2018).
- 47 Strack, M., and J. M. Waddington, 2008: Spatiotemporal variability in peatland subsurface methane
48 dynamics. *J. Geophys. Res. Biogeosciences*, **113**, doi:10.1029/2007JG000472.
49 <https://doi.org/10.1029/2007JG000472>.
- 50 Strandberg, G., and E. Kjellström, 2018: Climate Impacts from Afforestation and Deforestation in Europe.
51 *Earth Interact.*, **23**, 1–27, doi:10.1175/ei-d-17-0033.1.
- 52 Strefler, J., N. Bauer, E. Krieglner, A. Popp, A. Giannousakis, and O. Edenhofer, 2018: Between Scylla and
53 Charybdis : Delayed mitigation narrows the passage between large-scale CDR and high costs. *Environ.*
54 *Res. Lett.*, **13**, 44015, doi:10.1088/1748-9326/aab2ba.
- 55 Strengers, B. J., and Coauthors, 2010: Assessing 20th century climate-vegetation feedbacks of land-use
56 change and natural vegetation dynamics in a fully coupled vegetation-climate model. *Int. J. Climatol.*,

- 1 **30**, 2055–2065, doi:10.1002/joc.2132.
- 2 Sulman, Benjamin N., Elena Shevliakova, Edward R. Brzostek, Stephanie N. Kivlin, Sergey Malyshev,
3 Duncan NL Menge, and X. Z., 2019: Diverse mycorrhizal associations enhance terrestrial C storage in
4 a global model. *Global Biogeochem. Cycles*, doi:doi.org/10.1029/2018GB005973.
- 5 Sulman, B. N., R. P. Phillips, A. C. Oishi, E. Shevliakova, and S. W. Pacala, 2014a: Microbe-driven turnover
6 offsets mineral-mediated storage of soil carbon under elevated CO₂. *Nat. Clim. Chang.*, **4**, 1099–1102,
7 doi:10.1038/nclimate2436. <http://dx.doi.org/10.1038/nclimate2436>.
- 8 Sulman, B. N., R. P. Phillips, A. C. Oishi, E. Shevliakova, and S. W. Pacala, 2014b: Microbe-driven
9 turnover offsets mineral-mediated storage of soil carbon under elevated CO₂. *Nat. Clim. Chang.*, **4**, 1–31,
10 doi:10.1038/NCLIMATE2436.
- 11 Sulman, B. N., D. T. Roman, K. Yi, L. X. Wang, R. P. Phillips, and K. A. Novick, 2016: High atmospheric
12 demand for water can limit forest carbon uptake and transpiration as severely as dry soil. *Geophys. Res.
13 Lett.*, **43**, 9686–9695, doi:10.1002/2016gl069416.
- 14 Sulman, B. N., E. R. Brzostek, C. Medici, E. Shevliakova, D. N. L. Menge, and R. P. Phillips, 2017:
15 Feedbacks between plant N demand and rhizosphere priming depend on type of mycorrhizal
16 association. *Ecol. Lett.*, **20**, 1043–1053, doi:10.1111/ele.12802. <https://doi.org/10.1111/ele.12802>.
- 17 Sun, Y., X. Zhang, G. Ren, F. W. Zwiers, and T. Hu, 2016: Contribution of urbanization to warming in
18 China. *Nat. Clim. Chang.*, **6**, 706–709, doi:10.1038/nclimate2956.
- 19 Suni, T., and Coauthors, 2015: The significance of land-atmosphere interactions in the Earth system -
20 ILEAPS achievements and perspectives. *Anthropocene*, **12**, 69–84, doi:10.1016/j.ancene.2015.12.001.
- 21 Suntharalingam, P., E. Buitenhuis, C. Le Quéré, F. Dentener, C. Nevison, J. H. Butler, H. W. Bange, and G.
22 Forster, 2012: Quantifying the impact of anthropogenic nitrogen deposition on oceanic nitrous oxide.
23 *Geophys. Res. Lett.*, **39**, 1–7, doi:10.1029/2011GL050778.
- 24 Sussams, L. W., W. R. Sheate, and R. P. Eales, 2015: Green infrastructure as a climate change adaptation
25 policy intervention: Muddying the waters or clearing a path to a more secure future? *J. Environ.
26 Manage.*, **147**, 184–193, doi:10.1016/j.jenvman.2014.09.003.
- 27 Sutton, R. T., B. Dong, and J. M. Gregory, 2007: Land/sea warming ratio in response to climate change:
28 IPCC AR4 model results and comparison with observations. *Geophys. Res. Lett.*, **34**,
29 doi:10.1029/2006GL028164. <https://doi.org/10.1029/2006GL028164>.
- 30 Swann, A. L. S., I. Y. Fung, and J. C. H. Chiang, 2012a: Mid-latitude afforestation shifts general circulation
31 and tropical precipitation. *Proc. Natl. Acad. Sci.*, **109**, 712–716, doi:10.1073/pnas.1116706108.
32 <http://www.pnas.org/cgi/doi/10.1073/pnas.1116706108>.
- 33 —, —, and —, 2012b: Mid-latitude afforestation shifts general circulation and tropical precipitation.
34 *Proc. Natl. Acad. Sci.*, **109**, 712–716, doi:10.1073/pnas.1116706108.
35 <http://www.pnas.org/cgi/doi/10.1073/pnas.1116706108>.
- 36 Swann, A. L. S., F. M. Hoffman, C. D. Koven, and J. T. Randerson, 2016: Plant responses to increasing CO₂
37 reduce estimates of climate impacts on drought severity. *Proc. Natl. Acad. Sci.*, **113**, 10019–10024,
38 doi:10.1073/pnas.1604581113.
- 39 Tabari, H., and P. Willems, 2018: More prolonged droughts by the end of the century in the Middle East.
40 *Environ. Res. Lett.*, **13**, 104005, doi:10.1088/1748-9326/aae09c. [http://stacks.iop.org/1748-
41 9326/13/i=10/a=104005?key=crossref.051b189f636c30365c56a24f405cb9b7](http://stacks.iop.org/1748-9326/13/i=10/a=104005?key=crossref.051b189f636c30365c56a24f405cb9b7) (Accessed October 31,
42 2018).
- 43 Taheripour, F., and W. E. Tyner, 2013: Induced Land Use Emissions due to First and Second Generation
44 Biofuels and Uncertainty in Land Use Emission Factors. *Econ. Res. Int.*, doi:10.1155/2013/315787.
- 45 Taillardat, P., D. A. Friess, and M. Lupascu, 2018: Mangrove blue carbon strategies for climate change
46 mitigation are most effective at the national scale. *Biol. Lett.*, doi:10.1098/rsbl.2018.0251.
- 47 Talhelm, A. F., and Coauthors, 2014: Elevated carbon dioxide and ozone alter productivity and ecosystem
48 carbon content in northern temperate forests. *Glob. Chang. Biol.*, **20**, 2492–2504,
49 doi:10.1111/gcb.12564. <https://doi.org/10.1111/gcb.12564>.
- 50 Tan, Z. H., and Coauthors, 2017: Optimum air temperature for tropical forest photosynthesis: Mechanisms
51 involved and implications for climate warming. *Environ. Res. Lett.*, **12**, doi:10.1088/1748-9326/aa6f97.
- 52 Tang, B., X. Zhao, and W. Zhao, 2018: Local effects of forests on temperatures across Europe. *Remote
53 Sens.*, **10**, 1–24, doi:10.3390/rs10040529.
- 54 Tang, J., S. Luyssaert, A. D. Richardson, W. Kutsch, and I. A. Janssens, 2014: Steeper declines in forest
55 photosynthesis than respiration explain age-driven decreases in forest growth. *Proc. Natl. Acad. Sci.*,
56

- 1 **111**, 8856–8860, doi:10.1073/pnas.1320761111.
- 2 Tao, Y., F. Li, X. Liu, D. Zhao, X. Sun, and L. Xu, 2015: Variation in ecosystem services across an
3 urbanization gradient: A study of terrestrial carbon stocks from Changzhou, China. *Ecol. Modell.*, **318**,
4 210–216, doi:<https://doi.org/10.1016/j.ecolmodel.2015.04.027>.
- 5 Taraborrelli, D., M. G. Lawrence, J. N. Crowley, T. J. Dillon, S. Gromov, C. B. M. Groß, L. Vereecken, and
6 J. Lelieveld, 2012: Hydroxyl radical buffered by isoprene oxidation over tropical forests. *Nat. Geosci.*,
7 **5**, 190–193, doi:10.1038/ngeo1405.
- 8 Tate, K. R., 2015: Soil methane oxidation and land-use change – from process to mitigation. *Soil Biol.*
9 *Biochem.*, **80**, 260–272, doi:10.1016/j.soilbio.2014.10.010.
10 <http://dx.doi.org/10.1016/j.soilbio.2014.10.010>.
- 11 Tavoni, M., and Coauthors, 2015: Post-2020 climate agreements in the major economies assessed in the light
12 of global models. *Nat. Clim. Chang.*, **5**, 119–126, doi:10.1038/nclimate2475.
- 13 Taylor, C. M., and Coauthors, 2017: Frequency of extreme Sahelian storms tripled since 1982 in satellite
14 observations. *Nature*, **544**, 475–478, doi:10.1038/nature22069.
- 15 Taylor, L. L., and Coauthors, 2016: Enhanced weathering strategies for stabilizing climate and averting
16 ocean acidification. *Nat. Clim. Chang.*, **6**, 402–406, doi:10.1038/nclimate2882.
- 17 Teng, H., G. Branstator, G. A. Meehl, and W. M. Washington, 2016: Projected intensification of subseasonal
18 temperature variability and heat waves in the Great Plains. *Geophys. Res. Lett.*, **43**, 2165–2173,
19 doi:10.1002/2015GL067574. <http://doi.wiley.com/10.1002/2015GL067574> (Accessed October 28,
20 2018).
- 21 Tepley, A. J., E. Thomann, T. T. Veblen, G. L. W. Perry, A. Holz, J. Paritsis, T. Kitzberger, and K. J.
22 Anderson-Teixeira, 2018: Influences of fire–vegetation feedbacks and post-fire recovery rates on forest
23 landscape vulnerability to altered fire regimes. *J. Ecol.*, **106**, 1925–1940, doi:10.1111/1365-
24 2745.12950. <https://doi.org/10.1111/1365-2745.12950>.
- 25 Ter-Mikaelian, M. T., S. J. Colombo, and J. Chen, 2014: The Burning Question: Does Forest Bioenergy
26 Reduce Carbon Emissions? A Review of Common Misconceptions about Forest Carbon Accounting. *J.*
27 *For.*, **113**, 57–68, doi:10.5849/jof.14-016.
- 28 ———, and Coauthors, 2015: Carbon debt repayment or carbon sequestration parity? Lessons from a forest
29 bioenergy case study in Ontario, Canada. *GCB Bioenergy*, **7**, 704–716, doi:10.1111/gcbb.12198.
30 <http://dx.doi.org/10.1111/gcbb.12198>.
- 31 Teramoto, M., N. Liang, M. Takagi, J. Zeng, and J. Grace, 2016: Sustained acceleration of soil carbon
32 decomposition observed in a 6-year warming experiment in a warm-temperate forest in southern Japan.
33 *Sci. Rep.*, **6**, 35563. <https://doi.org/10.1038/srep35563>.
- 34 Terrer, C., S. Vicca, B. D. Stocker, B. A. Hungate, R. P. Phillips, P. B. Reich, A. C. Finzi, and I. C. Prentice,
35 2017: Ecosystem responses to elevated CO₂ governed by plant-soil interactions and the cost of
36 nitrogen acquisition. *New Phytol.*, **2**, 507–522, doi:10.1111/nph.14872.
37 <http://doi.wiley.com/10.1111/nph.14872>.
- 38 Teskey, R., T. Wertin, I. Bauweraerts, M. Ameye, M. A. McGuire, and K. Steppe, 2015: Responses of tree
39 species to heat waves and extreme heat events. *Plant Cell Environ.*, **38**, 1699–1712,
40 doi:10.1111/pce.12417. <http://doi.wiley.com/10.1111/pce.12417> (Accessed May 25, 2018).
- 41 Teuling, A. J., and Coauthors, 2010: Contrasting response of European forest and grassland energy exchange
42 to heatwaves. *Nat. Geosci.*, **3**, 722–727, doi:10.1038/ngeo950. <http://www.nature.com/articles/ngeo950>
43 (Accessed October 30, 2018).
- 44 Thackeray, C. W., and C. G. Fletcher, 2015: Snow albedo feedback: Current knowledge, importance,
45 outstanding issues and future directions. *Prog. Phys. Geogr.*, **40**, 392–408,
46 doi:10.1177/0309133315620999.
- 47 Thalman, R., and Coauthors, 2017: CCN activity and organic hygroscopicity of aerosols downwind of an
48 urban region in central Amazonia: Seasonal and diel variations and impact of anthropogenic emissions.
49 *Atmos. Chem. Phys.*, **17**, 11779–11801, doi:10.5194/acp-17-11779-2017.
- 50 Tharammal, T., G. Bala, D. Narayanappa, and R. Nemani, 2018: Potential roles of CO₂ fertilization,
51 nitrogen deposition, climate change, and land use and land cover change on the global terrestrial carbon
52 uptake in the twenty-first century. *Climate Dynamics*.
- 53 Thiery, W., E. L. Davin, D. M. Lawrence, A. L. Hirsch, M. Hauser, and S. I. Seneviratne, 2017: Present-day
54 irrigation mitigates heat extremes. *J. Geophys. Res.*, **122**, 1403–1422, doi:10.1002/2016JD025740.
55 <https://doi.org/10.1002/2016JD025740>.
- 56 Thompson, R. L., and Coauthors, 2014: TransCom N₂O model inter-comparison - Part 2: Atmospheric

- 1 inversion estimates of N₂O emissions. *Atmos. Chem. Phys.*, **14**, 6177–6194, doi:10.5194/acp-14-6177-
2 2014.
- 3 Thompson, V., N. J. Dunstone, A. A. Scaife, D. M. Smith, J. M. Slingo, S. Brown, and S. E. Belcher, 2017:
4 High risk of unprecedented UK rainfall in the current climate. *Nat. Commun.*, **8**, 107,
5 doi:10.1038/s41467-017-00275-3. <http://www.nature.com/articles/s41467-017-00275-3> (Accessed May
6 24, 2018).
- 7 Tian, H., and Coauthors, 2015: Global patterns and controls of soil organic carbon dynamics as simulated by
8 multiple terrestrial biosphere models: Current status and future directions. *Global Biogeochem. Cycles*,
9 **29**, 775–792, doi:10.1002/2014GB005021. [http://www.fao.org/forest-resources-assessment/current-
11 assessment/en/](http://www.fao.org/forest-resources-assessment/current-
10 assessment/en/).
- 12 ———, and Coauthors, 2016: The terrestrial biosphere as a net source of greenhouse gases to the atmosphere.
13 *Nature*, **531**, 225–228, doi:10.1038/nature16946.
14 <http://www.nature.com/doi/finder/10.1038/nature16946>.
- 15 ———, and Coauthors, 2018: The Global N₂O Model Intercomparison Project. *Bull. Am. Meteorol. Soc.*, **99**,
16 1231–1251, doi:10.1175/BAMS-D-17-0212.1. <http://www.ncbi.nlm.nih.gov/pubmed/18932145>.
- 17 Tietjen, B., and Coauthors, 2017: Climate change-induced vegetation shifts lead to more ecological droughts
18 despite projected rainfall increases in many global temperate drylands. *Glob. Chang. Biol.*, **23**, 2743–
19 2754, doi:10.1111/gcb.13598.
- 20 Tifafi, M., B. Guenet, and C. Hatté, 2018: Large Differences in Global and Regional Total Soil Carbon Stock
21 Estimates Based on SoilGrids, HWSD, and NCSCD: Intercomparison and Evaluation Based on Field
22 Data From USA, England, Wales, and France. *Global Biogeochem. Cycles*, **32**, 42–56,
23 doi:10.1002/2017GB005678. <http://doi.wiley.com/10.1002/2017GB005678>.
- 24 Tilman, D., and M. Clark, 2014: Global diets link environmental sustainability and human health. *Nature*,
25 **515**, 518.
- 26 Tjoelker, M. G., 2018: The role of thermal acclimation of plant respiration under climate warming: Putting
27 the brakes on a runaway train? *Plant Cell Environ.*, **41**, 501–503, doi:10.1111/pce.13126.
28 <http://doi.wiley.com/10.1111/pce.13126>.
- 29 Tobarra, M. A., L. A. López, M. A. Cadarso, N. Gómez, and I. Cazarro, 2018: Is Seasonal Households’
30 Consumption Good for the Nexus Carbon/Water Footprint? The Spanish Fruits and Vegetables Case.
31 *Environ. Sci. Technol.*, **52**, 12066–12077, doi:10.1021/acs.est.8b00221.
32 <https://doi.org/10.1021/acs.est.8b00221>.
- 33 Todd-Brown, K. E. O., J. T. Randerson, W. M. Post, F. M. Hoffman, C. Tarnocai, E. A. G. Schuur, and S. D.
34 Allison, 2013: Causes of variation in soil carbon simulations from CMIP5 Earth system models and
35 comparison with observations. *Biogeosciences*, **10**, 1717–1736, doi:10.5194/bg-10-1717-2013.
36 <http://www.biogeosciences.net/10/1717/2013/>.
- 37 Tölle, M. H., S. Engler, and H. J. Panitz, 2017: Impact of abrupt land cover changes by tropical deforestation
38 on Southeast Asian climate and agriculture. *J. Clim.*, **30**, 2587–2600, doi:10.1175/JCLI-D-16-0131.1.
- 39 Torres-Valcárcel, Á. R., J. Harbor, A. L. Torres-Valcárcel, and C. J. González-Avilés, 2015: Historical
40 differences in temperature between urban and non-urban areas in Puerto Rico. *Int. J. Climatol.*, **35**,
41 1648–1661, doi:10.1002/joc.4083.
- 42 Toyoda, S., N. Kuroki, N. Yoshida, K. Ishijima, Y. Tohjima, and T. Machida, 2013: Decadal time series of
43 tropospheric abundance of N₂O isotopomers and isotopologues in the Northern Hemisphere obtained
44 by the long-term observation at Hateruma Island, Japan. *J. Geophys. Res. Atmos.*, **118**, 3369–3381,
45 doi:10.1002/jgrd.50221. <http://doi.wiley.com/10.1002/jgrd.50221>.
- 46 Trenberth, K. E., 2012: Framing the way to relate climate extremes to climate change. *Clim. Change*, **115**,
47 283–290, doi:10.1007/s10584-012-0441-5.
- 48 ———, A. Dai, G. Van Der Schrier, P. D. Jones, J. Barichivich, K. R. Briffa, and J. Sheffield, 2014: Global
49 warming and changes in drought. *Nat. Clim. Chang.*, **4**, 17–22, doi:10.1038/nclimate2067.
50 <http://www.nature.com/articles/nclimate2067> (Accessed May 7, 2018).
- 51 ———, J. T. Fasullo, and T. G. Shepherd, 2015: Attribution of climate extreme events. *Nat. Clim. Chang.*, **5**,
52 725–730, doi:10.1038/nclimate2657.
- 53 Tribouillois, H., J. Constantin, and E. Justes, 2018: Cover crops mitigate direct greenhouse gases balance but
54 reduce drainage under climate change scenarios in temperate climate with dry summers. *Glob. Chang.*
55 *Biol.*, **24**, 2513–2529, doi:10.1111/gcb.14091.
- 56 Trumbore, S., P. Brando, and H. Hartmann, 2015: Forest health and global change. *Science (80-.)*, **349**,
814–818, doi:10.1126/science.aac6759. <http://www.ncbi.nlm.nih.gov/pubmed/26293952> (Accessed

- 1 March 24, 2019).
- 2 Tsigaridis, K., and Coauthors, 2014a: The AeroCom evaluation and intercomparison of organic aerosol in
3 global models. *Atmos. Chem. Phys.*, **14**, 10845–10895, doi:10.5194/acp-14-10845-2014.
- 4 —, and Coauthors, 2014b: The AeroCom evaluation and intercomparison of organic aerosol in global
5 models. *Atmos. Chem. Phys.*, **14**, 10845–10895, doi:10.5194/acp-14-10845-2014. <https://www.atmos-chem-phys.net/14/10845/2014/>.
- 6
- 7 Tubiello, F. N., M. Salvatore, S. Rossi, A. Ferrara, N. Fitton, and P. Smith, 2013: The FAOSTAT database
8 of greenhouse gas emissions from agriculture. *Environ. Res. Lett.*, **8**, 15009, doi:10.1088/1748-
9 9326/8/1/015009. <http://stacks.iop.org/1748-9326/8/i=1/a=015009?key=crossref.fa67310f793955d24ab78eafdc220a33>.
- 10
- 11 Tubiello, F. N., and Coauthors, 2014: *Agriculture, Forestry and Other Land Use Emissions by Sources and*
12 *Removals by Sinks*. 4-89 pp. <http://www.fao.org/docrep/019/i3671e/i3671e.pdf>.
- 13 Tubiello, F. N., and Coauthors, 2015: The Contribution of Agriculture, Forestry and other Land Use
14 activities to Global Warming, 1990-2012. *Glob. Chang. Biol.*, **21**, 2655–2660, doi:10.1111/gcb.12865.
- 15 Tuittila, E.-S., V.-M. Komulainen, H. Vasander, H. Nykänen, P. J. Martikainen, and J. Laine, 2000: Methane
16 dynamics of a restored cut-away peatland. *Glob. Chang. Biol.*, **6**, 569–581, doi:10.1046/j.1365-
17 2486.2000.00341.x. <http://doi.wiley.com/10.1046/j.1365-2486.2000.00341.x>.
- 18 Turetsky, M. R., B. Benscoter, S. Page, G. Rein, G. R. van der Werf, and A. Watts, 2014: Global
19 vulnerability of peatlands to fire and carbon loss. *Nat. Geosci.*, **8**, 11.
20 <http://dx.doi.org/10.1038/ngeo2325>.
- 21 Turner, A. J., C. Frankenberg, P. O. Wennberg, and D. J. Jacob, 2017: Ambiguity in the causes for decadal
22 trends in atmospheric methane and hydroxyl. *Proc. Natl. Acad. Sci.*, **114**, 5367–5372,
23 doi:10.1073/pnas.1616020114. <http://www.pnas.org/lookup/doi/10.1073/pnas.1616020114>.
- 24 Turner, B. L., T. Brenes-Arguedas, and R. Condit, 2018a: Pervasive phosphorus limitation of tree species but
25 not communities in tropical forests. *Nature*, **555**, 367–370, doi:10.1038/nature25789.
26 <http://dx.doi.org/10.1038/nature25789>.
- 27 Turner, P. A., C. B. Field, D. B. Lobell, D. L. Sanchez, and K. J. Mach, 2018b: Unprecedented rates of land-
28 use transformation in modelled climate change mitigation pathways. *Nat. Sustain.*, **1**, 240–245,
29 doi:10.1038/s41893-018-0063-7.
- 30 —, —, —, —, and —, 2018c: Unprecedented rates of land-use transformation in modelled
31 climate change mitigation pathways. *Nat. Sustain.*, **1**, 240–245, doi:10.1038/s41893-018-0063-7.
- 32 Tuttle, S., and G. Salvucci, 2016: Atmospheric science: Empirical evidence of contrasting soil moisture-
33 precipitation feedbacks across the United States. *Science (80-.)*, **352**, 825–828,
34 doi:10.1126/science.aaa7185. <http://science.sciencemag.org/content/352/6287/825>.
- 35 Tyukavina, A., and Coauthors, 2015: Aboveground carbon loss in natural and managed tropical forests from
36 2000 to 2012. *Environ. Res. Lett.*, **10**, 74002–74002, doi:10.1088/1748-9326/10/7/074002.
37 <http://iopscience.iop.org/1748-9326/10/7/074002>.
- 38 Tyukavina, A., M. C. Hansen, P. V. Potapov, S. V. Stehman, K. Smith-Rodriguez, C. Okpa, and R. Aguilar,
39 2017: Types and rates of forest disturbance in Brazilian Legal Amazon, 2000–2013. *Sci. Adv.*, **3**,
40 e1601047, doi:10.1126/sciadv.1601047.
41 <http://advances.sciencemag.org/lookup/doi/10.1126/sciadv.1601047>.
- 42 Ukkola, A. M., A. J. Pitman, M. G. Donat, M. G. De Kauwe, and O. Angéllil, 2018: Evaluating the
43 Contribution of Land-Atmosphere Coupling to Heat Extremes in CMIP5 Models. *Geophys. Res. Lett.*,
44 **45**, 9003–9012, doi:10.1029/2018GL079102.
- 45 Ummenhofer, C. C., and G. A. Meehl, 2017: Extreme weather and climate events with ecological relevance:
46 a review. *Philos. Trans. R. Soc. B Biol. Sci.*, **372**, 20160135, doi:10.1098/rstb.2016.0135.
- 47 UNEP, P. D. Roberts, G. B. Stewart, and A. S. Pullin, 2017: *The Emissions Gap Report 2017 - A UN*
48 *Environment Synthesis Report*. 409-423 pp.
- 49 Unger, N., 2013: Isoprene emission variability through the twentieth century. *J. Geophys. Res. Atmos.*, **118**,
50 13,606–613,613, doi:10.1002/2013JD020978.
51 <https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1002/2013JD020978>.
- 52 Unger, N., 2014a: On the role of plant volatiles in anthropogenic global climate change. *Geophys. Res. Lett.*,
53 **41**, 8563–8569, doi:10.1002/2014GL061616. <http://dx.doi.org/10.1002/2014gl061616>.
- 54 —, 2014b: Human land-use-driven reduction of forest volatiles cools global climate. *Nat. Clim. Chang.*, **4**,
55 907–910, doi:10.1038/nclimate2347.
- 56 Unger, N., X. Yue, and K. L. Harper, 2017: Aerosol climate change effects on land ecosystem services.

- 1 *Faraday Discuss.*, **200**, 121–142, doi:10.1039/c7fd00033b.
- 2 United Nations, 2017: *World Population Prospects: The 2017 Revision, Key Findings and Advance Tables*.
- 3 ESA/P/WP/248 pp.
- 4 Urban, M. C., 2015a: Accelerating extinction risk from climate change. *Science (80-.)*, **348**, 571–573,
- 5 doi:10.1126/science.aaa4984.
- 6 ———, 2015b: Accelerating extinction risk from climate change. *Science (80-.)*, **348**, 571–573,
- 7 doi:10.1126/science.aaa4984.
- 8 Urbanski, S. P., 2014: Wildland fire emissions, carbon, and climate: Emission factors. *For. Ecol. Manage.*,
- 9 **317**, 51–60, doi:https://doi.org/10.1016/j.foreco.2013.05.045.
- 10 <http://www.sciencedirect.com/science/article/pii/S0378112713003447>.
- 11 USEPA, 2012: *Global Anthropogenic Non-CO₂ Greenhouse Gas Emissions : 1990 - 2030*. Washington
- 12 D.C., 176 pp.
- 13 ———, 2013: *Global Mitigation of Non-CO₂ Greenhouse Gases: 2010-2030*. United States Environmental
- 14 Protection Agency, Washington DC, 410 pp.
- 15 http://www.epa.gov/climatechange/Downloads/EPAactivities/MAC_Report_2013.pdf.
- 16 Valade, A., C. Magand, S. Luyssaert, and V. Bellassen, 2017: Sustaining the sequestration efficiency of the
- 17 European forest sector. *For. Ecol. Manage.*, **405**, 44–55, doi:10.1016/j.foreco.2017.09.009.
- 18 <http://europemc.org/abstract/AGR/IND605824755>.
- 19 ———, S. Luyssaert, P. Vallet, S. Njakou Djomo, I. Jesus Van Der Kellen, and V. Bellassen, 2018: Carbon
- 20 costs and benefits of France’s biomass energy production targets. *Carbon Balance Manag.*,
- 21 doi:10.1186/s13021-018-0113-5.
- 22 Valin, H., D. Peters, van den B. M, S. Frank, P. Havlik, N. Forsell, C. Hamelinck, and ., 2015: *The land use*
- 23 *change impact of biofuels consumed in the EU: Quantification of area and greenhouse gas impacts*.
- 24 ECOFYS Netherlands B.V., Utrecht, Netherlands,.
- 25 Vanderwel, M. C., and Coauthors, 2015: Global convergence in leaf respiration from estimates of thermal
- 26 acclimation across time and space. *New Phytol.*, **207**, 1026–1037, doi:10.1111/nph.13417.
- 27 Vanselow-Algan, M., S. R. Schmidt, M. Greven, C. Fiencke, L. Kutzbach, and E.-M. Pfeiffer, 2015: High
- 28 methane emissions dominated annual greenhouse gas balances 30 years after bog rewetting.
- 29 *Biogeosciences*, **12**, 4361–4371, doi:10.5194/bg-12-4361-2015.
- 30 <https://www.biogeosciences.net/12/4361/2015/>.
- 31 Vaughan, N. E., C. Gough, S. Mander, E. W. Littleton, A. Welfle, D. E. H. J. Gernaat, and D. P. van Vuuren,
- 32 2018: Evaluating the use of biomass energy with carbon capture and storage in low emission scenarios.
- 33 *Environ. Res. Lett.*, **13**, 44014. <http://stacks.iop.org/1748-9326/13/i=4/a=044014>.
- 34 Veraverbeke, S., I. Gitas, T. Katagis, A. Polychronaki, B. Somers, and R. Goossens, 2012: Assessing post-
- 35 fire vegetation recovery using red-near infrared vegetation indices: Accounting for background and
- 36 vegetation variability. *ISPRS J. Photogramm. Remote Sens.*, **68**, 28–39,
- 37 doi:10.1016/j.isprsjprs.2011.12.007.
- 38 Veraverbeke, S., B. M. Rogers, M. L. Goulden, R. R. Jandt, C. E. Miller, E. B. Wiggins, and J. T.
- 39 Randerson, 2017: Lightning as a major driver of recent large fire years in North American boreal
- 40 forests. *Nat. Clim. Chang.*, **7**, 529, doi:10.1038/nclimate3329. <http://dx.doi.org/10.1038/nclimate3329>.
- 41 Verbyla, D., 2008: The greening and browning of Alaska based on 1982-2003 satellite data. *Glob. Ecol.*
- 42 *Biogeogr.*, **17**, 547–555, doi:10.1111/j.1466-8238.2008.00396.x. [http://doi.wiley.com/10.1111/j.1466-](http://doi.wiley.com/10.1111/j.1466-8238.2008.00396.x)
- 43 [8238.2008.00396.x](http://doi.wiley.com/10.1111/j.1466-8238.2008.00396.x) (Accessed March 12, 2019).
- 44 Verheijen, F. G. A., S. Jeffery, M. Van Der Velde, V. Penížek, M. Beland, A. C. Bastos, and J. J. Keizer,
- 45 2013: Reductions in soil surface albedo as a function of biochar application rate: Implications for
- 46 global radiative forcing. *Environ. Res. Lett.*, doi:10.1088/1748-9326/8/4/044008.
- 47 Vermeulen, S. J., B. M. Campbell, and J. S. I. Ingram, 2012a: Climate Change and Food Systems. *Annu.*
- 48 *Rev. Environ. Resour.*, **37**, 195–222, doi:10.1146/annurev-environ-020411-130608.
- 49 Vermeulen, S. J., B. M. Campbell, and J. S. I. I. Ingram, 2012b: Climate Change and Food Systems. *Annu.*
- 50 *Rev. Environ. Resour.*, **37**, 195–222, doi:10.1146/annurev-environ-020411-130608.
- 51 <http://www.annualreviews.org/doi/10.1146/annurev-environ-020411-130608> (Accessed September 29,
- 52 2018).
- 53 Veselovskii, I., and Coauthors, 2016: Retrieval of optical and physical properties of African dust from
- 54 multiwavelength Raman lidar measurements during the SHADOW campaign in Senegal. *Atmos. Chem.*
- 55 *Phys.*, **16**, 7013–7028, doi:10.5194/acp-16-7013-2016.
- 56 Vicente-Serrano, S. M., and Coauthors, 2014: Evidence of increasing drought severity caused by temperature

- 1 rise in southern Europe. *Environ. Res. Lett.*, **9**, 44001, doi:10.1088/1748-9326/9/4/044001.
2 <http://stacks.iop.org/1748-9326/9/i=4/a=044001?key=crossref.a4db08a1ad24a00be1d79dd69814b0e5>
3 (Accessed October 29, 2018).
- 4 Vogel, E., and R. Meyer, 2018: Climate Change, Climate Extremes, and Global Food Production—
5 Adaptation in the Agricultural Sector. *Resilience*, 31–49, doi:10.1016/B978-0-12-811891-7.00003-7.
6 <http://linkinghub.elsevier.com/retrieve/pii/B9780128118917000037> (Accessed May 25, 2018).
- 7 ———, M. Donat, L. V Alexander, M. Meinshausen, D. Ray, D. Karoly, N. Meinshausen, and K. Frieler,
8 2019: The effects of climate extremes on global agricultural yields. *Environ. Res. Lett.*,
9 doi:10.1088/1748-9326/ab154b. <http://iopscience.iop.org/article/10.1088/1748-9326/ab154b> (Accessed
10 April 12, 2019).
- 11 Vogel, M. M., R. Orth, F. Cheruy, S. Hagemann, R. Lorenz, B. J. J. M. van den Hurk, and S. I. Seneviratne,
12 2017: Regional amplification of projected changes in extreme temperatures strongly controlled by soil
13 moisture-temperature feedbacks. *Geophys. Res. Lett.*, **44**, 1511–1519, doi:10.1002/2016GL071235.
- 14 Voldoire, A., 2006: Quantifying the impact of future land-use changes against increases in GHG
15 concentrations. *Geophys. Res. Lett.*, **33**, 2–5, doi:10.1029/2005GL024354.
- 16 De Vrese, P., S. Hagemann, and M. Claussen, 2016a: Asian irrigation, African rain: Remote impacts of
17 irrigation. *Geophys. Res. Lett.*, **43**, 3737–3745, doi:10.1002/2016GL068146.
- 18 ———, ———, and ———, 2016b: Asian irrigation, African rain: Remote impacts of irrigation. *Geophys. Res.*
19 *Lett.*, **43**, 3737–3745, doi:10.1002/2016GL068146.
- 20 van Vuuren, D. P., and Coauthors, 2018: Alternative pathways to the 1.5 °C target reduce the need for
21 negative emission technologies. *Nat. Clim. Chang.*, **8**, 1–7, doi:10.1038/s41558-018-0119-8.
- 22 Wäldchen, J., E.-D. Schulze, I. Schöning, M. Schrumpf, and C. Sierra, 2013: The influence of changes in
23 forest management over the past 200years on present soil organic carbon stocks. *For. Ecol. Manage.*,
24 **289**, 243–254, doi:https://doi.org/10.1016/j.foreco.2012.10.014.
- 25 Wallace, C. J., and M. Joshi, 2018: Comparison of land-ocean warming ratios in updated observed records
26 and CMIP5 climate models. *Environ. Res. Lett.*, **13**, doi:10.1088/1748-9326/aae46f.
- 27 Wang, B., and Q. Ding, 2008: Global monsoon: Dominant mode of annual variation in the tropics. *Dyn.*
28 *Atmos. Ocean.*, **44**, 165–183, doi:10.1016/j.dynatmoce.2007.05.002.
- 29 Wang, G. Q., and Coauthors, 2013: Simulating the Impact of Climate Change on Runoff in a Typical River
30 Catchment of the Loess Plateau, China. *J. Hydrometeorol.*, **14**, 1553–1561, doi:10.1175/JHM-D-12-
31 081.1. <https://doi.org/10.1175/JHM-D-12-081.1>.
- 32 Wang, J., and Coauthors, 2016a: Amazon boundary layer aerosol concentration sustained by vertical
33 transport during rainfall. *Nature*, **539**, 416–419, doi:10.1038/nature19819.
- 34 Wang, J., B. Huang, D. Fu, P. M. Atkinson, and X. Zhang, 2016b: Response of urban heat island to future
35 urban expansion over the Beijing-Tianjin-Hebei metropolitan area. *Appl. Geogr.*, **70**, 26–36,
36 doi:10.1016/j.apgeog.2016.02.010.
37 <http://www.sciencedirect.com/science/article/pii/S014362281630011X>.
- 38 Wang, J., Z. Yan, X. W. Quan, and J. Feng, 2017a: Urban warming in the 2013 summer heat wave in eastern
39 China. *Clim. Dyn.*, **48**, 3015–3033, doi:10.1007/s00382-016-3248-7.
- 40 Wang, M., and Coauthors, 2017b: On the Long-Term Hydroclimatic Sustainability of Perennial Bioenergy
41 Crop Expansion over the United States. *J. Clim.*, **30**, 2535–2557, doi:10.1175/jcli-d-16-0610.1.
42 <http://journals.ametsoc.org/doi/abs/10.1175/JCLI-D-16-0610.1>.
- 43 Wang, R., and Coauthors, 2014a: Trend in global black carbon emissions from 1960 to 2007. *Environ. Sci.*
44 *Technol.*, **48**, 6780–6787, doi:10.1021/es5021422.
- 45 ———, and Coauthors, 2017c: Global forest carbon uptake due to nitrogen and phosphorus deposition from
46 1850 to 2100. *Glob. Chang. Biol.*, **23**, 4854–4872, doi:10.1111/gcb.13766.
- 47 Wang, W., A. T. Evan, C. Flamant, and C. Lavaysse, 2015: On the decadal scale correlation between African
48 dust and Sahel rainfall: The role of Saharan heat low-forced winds. *Sci. Adv.*, **1**, p.e1500646,
49 doi:10.1126/sciadv.1500646.
- 50 Wang, W., and Coauthors, 2016c: Evaluation of air-soil temperature relationships simulated by land surface
51 models during winter across the permafrost region. *Cryosphere*, **10**, 1721–1737, doi:10.5194/tc-10-
52 1721-2016.
- 53 Wang, X., and Coauthors, 2016d: Deriving brown carbon from multiwavelength absorption measurements:
54 Method and application to AERONET and Aethalometer observations. *Atmos. Chem. Phys.*, **16**,
55 12733–12752, doi:10.5194/acp-16-12733-2016.
- 56 Wang, Y., X. Yan, and Z. Wang, 2014b: The biogeophysical effects of extreme afforestation in modeling

- 1 future climate. *Theor. Appl. Climatol.*, **118**, 511–521, doi:10.1007/s00704-013-1085-8.
- 2 Wang, Z., and Coauthors, 2017d: Human-induced erosion has offset one-third of carbon emissions from land
3 cover change. *Nat. Clim. Chang.*, **7**, 345–349, doi:10.1038/nclimate3263.
4 <http://dx.doi.org/10.1038/nclimate3263>.
- 5 Wang, Z., J. Li, C. Lai, Z. Zeng, R. Zhong, X. Chen, X. Zhou, and M. Wang, 2017e: Does drought in China
6 show a significant decreasing trend from 1961 to 2009? *Sci. Total Environ.*, **579**, 314–324,
7 doi:10.1016/j.scitotenv.2016.11.098.
8 <https://www.sciencedirect.com/science/article/pii/S0048969716325372?via%3Dihub> (Accessed
9 October 29, 2018).
- 10 Ward, D. S., N. M. Mahowald, and S. Kloster, 2014: Potential climate forcing of land use and land cover
11 change. *Atmos. Chem. Phys.*, **14**, 12701–12724, doi:10.5194/acp-14-12701-2014.
- 12 Ward, K., S. Lauf, B. Kleinschmit, and W. Endlicher, 2016: Heat waves and urban heat islands in Europe: A
13 review of relevant drivers. *Sci. Total Environ.*, **569–570**, 527–539, doi:10.1016/j.scitotenv.2016.06.119.
- 14 Warszawski, L., and Coauthors, 2013: A multi-model analysis of risk of ecosystem shifts under climate
15 change. *Environ. Res. Lett.*, **8**, 44018, doi:10.1088/1748-9326/8/4/044018. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/8/i=4/a=044018?key=crossref.2d722efbaf6d3e455e38f2c68888d873)
16 [9326/8/i=4/a=044018?key=crossref.2d722efbaf6d3e455e38f2c68888d873](http://stacks.iop.org/1748-9326/8/i=4/a=044018?key=crossref.2d722efbaf6d3e455e38f2c68888d873) (Accessed May 7, 2018).
- 17 Watts, N., W. N. Adger, and P. Agnolucci, 2015: Health and climate change: Policy responses to protect
18 public health. *Environnement, Risques et Sante*, **14**, 466–468, doi:10.1016/S0140-6736(15)60854-6.
- 19 Wehner, M. F., 2013: Very extreme seasonal precipitation in the NARCCAP ensemble: Model performance
20 and projections. *Clim. Dyn.*, **40**, 59–80, doi:10.1007/s00382-012-1393-1.
21 <http://link.springer.com/10.1007/s00382-012-1393-1> (Accessed October 30, 2018).
- 22 Wells, K. C., and Coauthors, 2018: Top-down constraints on global N₂O emissions at optimal resolution:
23 Application of a new dimension reduction technique. *Atmos. Chem. Phys.*, **18**, 735–756,
24 doi:10.5194/acp-18-735-2018.
- 25 Wendt, J. W., and S. Hauser, 2013: An equivalent soil mass procedure for monitoring soil organic carbon in
26 multiple soil layers. *Eur. J. Soil Sci.*, **64**, 58–65, doi:10.1111/ejss.12002.
- 27 Wenz, L., A. Levermann, and M. Auffhammer, 2017: North–south polarization of European electricity
28 consumption under future warming. *Proc. Natl. Acad. Sci.*, **114**, E7910–E7918,
29 doi:10.1073/pnas.1704339114.
- 30 van der Werf, G. R., C. P. Meyer, and A. P., 2013: *Reducing N₂O emissions from biomass burning in
31 landscape fires and household stoves Drawing Down N₂O to Protect Climate and the Ozone Layer*.
- 32 van der Werf, G. R., and Coauthors, 2017: Global fire emissions estimates during
33 1997&ndash;2015. *Earth Syst. Sci. Data Discuss.*, 1–43, doi:10.5194/essd-2016-62.
34 <http://www.earth-syst-sci-data-discuss.net/essd-2016-62/>.
- 35 Van Der Werf, G. R., and Coauthors, 2010: Global fire emissions and the contribution of deforestation,
36 savanna, forest, agricultural, and peat fires (1997–2009). *Atmos. Chem. Phys.*, **10**, 11707–11735,
37 doi:10.5194/acp-10-11707-2010. <http://www.atmos-chem-phys-discuss.net/6/3175/2006/>.
- 38 Werth, D., 2002: The local and global effects of Amazon deforestation. *J. Geophys. Res.*, **107**, 8087,
39 doi:10.1029/2001jd000717. <http://doi.wiley.com/10.1029/2001JD000717> (Accessed January 10, 2018).
- 40 ———, 2005: The local and global effects of African deforestation. *Geophys. Res. Lett.*, **32**, 1–4,
41 doi:10.1029/2005GL022969.
- 42 ———, and R. Avissar, 2005: The local and global effects of Southeast Asian deforestation. *Geophys. Res.
43 Lett.*, **32**, 1–4, doi:10.1029/2005GL022970.
- 44 West, P. C., G. T. Narisma, C. C. Barford, C. J. Kucharik, and J. a. Foley, 2011: An alternative approach for
45 quantifying climate regulation by ecosystems. *Front. Ecol. Environ.*, **9**, 126–133, doi:10.1890/090015.
- 46 West, T. O., G. Marland, A. W. King, W. M. Post, A. K. Jain, and K. Andrasko, 2004: Carbon management
47 response curves: Estimates of temporal soil carbon dynamics. *Environ. Manage.*, **33**, 507–518,
48 doi:10.1007/s00267-003-9108-3.
- 49 Westervelt, D. M., L. W. Horowitz, V. Naik, J. C. Golaz, and D. L. Mauzerall, 2015: Radiative forcing and
50 climate response to projected 21st century aerosol decreases. *Atmos. Chem. Phys.*, **15**, 12681–12703,
51 doi:10.5194/acp-15-12681-2015. <https://www.atmos-chem-phys.net/15/12681/2015/>.
- 52 Westra, S., L. V. Alexander, and F. W. Zwiers, 2013: Global increasing trends in annual maximum daily
53 precipitation. *J. Clim.*, **26**, 3904–3918, doi:10.1175/JCLI-D-12-00502.1.
54 <http://journals.ametsoc.org/doi/abs/10.1175/JCLI-D-12-00502.1> (Accessed May 21, 2018).
- 55 Wheeler, T., and J. Von Braun, 2013: Climate change impacts on global food security. *Science (80-.)*, **341**,
56 508–513, doi:10.1126/science.1239402.

- 1 Whitehead, J. D., and Coauthors, 2016: Biogenic cloud nuclei in the central Amazon during the transition
2 from wet to dry season. *Atmos. Chem. Phys.*, **16**, 9727–9743, doi:10.5194/acp-16-9727-2016.
- 3 Wicke, B., P. Verweij, H. van Meijl, D. P. van Vuuren, and A. P. C. Faaij, 2012: Indirect land use change:
4 review of existing models and strategies for mitigation. *Biofuels*, **3**, 87–100, doi:10.4155/bfs.11.154.
5 <https://doi.org/10.4155/bfs.11.154>.
- 6 Wickham, J., T. G. Wade, and K. H. Riitters, 2014: An isoline separating relatively warm from relatively
7 cool wintertime forest surface temperatures for the southeastern United States. *Glob. Planet. Change*,
8 **120**, 46–53, doi:10.1016/j.gloplacha.2014.05.012. <http://dx.doi.org/10.1016/j.gloplacha.2014.05.012>.
- 9 Wickham, J. D., T. G. Wade, and K. H. Riitters, 2013: Empirical analysis of the influence of forest extent on
10 annual and seasonal surface temperatures for the continental United States. *Glob. Ecol. Biogeogr.*, **22**,
11 620–629, doi:10.1111/geb.12013.
- 12 Wieder, William R., Bonan, Gordon B., Allison, S. D., 2013: Global Soil Carbon Projections are Improved
13 by Modeling Microbial Processes. *Nat. Clim. Chang.*, **3**, 909. <http://dx.doi.org/10.1038/nclimate1951>.
- 14 Wieder, W. R., C. C. Cleveland, W. K. Smith, and K. Todd-Brown, 2015: Future productivity and carbon
15 storage limited by terrestrial nutrient availability. *Nat. Geosci.*, **8**, 441–444, doi:10.1038/NNGEO2413.
- 16 Wieder, W. R., M. D. Hartman, B. N. Sulman, Y. P. Wang, C. D. Koven, and G. B. Bonan, 2018: Carbon
17 cycle confidence and uncertainty: Exploring variation among soil biogeochemical models. *Glob.*
18 *Chang. Biol.*, **24**, 1563–1579, doi:10.1111/gcb.13979.
- 19 Wienert, U., and W. Kuttler, 2005: The dependence of the urban heat island intensity on latitude - A
20 statistical approach. *Meteorol. Zeitschrift*, **14**, 677–686, doi:10.1127/0941-2948/2005/0069.
- 21 Wilcox, L. J., and Coauthors, 2018: Multiple perspectives on the attribution of the extreme European
22 summer of 2012 to climate change. *Clim. Dyn.*, **50**, 3537–3555, doi:10.1007/s00382-017-3822-7.
23 <http://link.springer.com/10.1007/s00382-017-3822-7> (Accessed October 29, 2018).
- 24 Wilhelm, J. A., and R. G. Smith, 2018: Ecosystem services and land sparing potential of urban and peri-
25 urban agriculture: A review. *Renew. Agric. Food Syst.*, **33**, 481–494, doi:DOI:
26 10.1017/S1742170517000205. [https://www.cambridge.org/core/article/ecosystem-services-and-land-
27 sparing-potential-of-urban-and-periurban-agriculture-a-
28 review/22C39E215157FA9C3E7ABCB11DE9C742](https://www.cambridge.org/core/article/ecosystem-services-and-land-sparing-potential-of-urban-and-periurban-agriculture-a-review/22C39E215157FA9C3E7ABCB11DE9C742).
- 29 Wilhelm, M., E. L. Davin, and S. I. Seneviratne, 2015: Journal of Geophysical Research : Atmospheres
30 mitigation : An Earth system model sensitivity study. 2612–2623,
31 doi:10.1002/2014JD022293.Received.
- 32 Williams, A. P., R. Seager, J. T. Abatzoglou, B. I. Cook, J. E. Smerdon, and E. R. Cook, 2015: Contribution
33 of anthropogenic warming to California drought during 2012–2014. *Geophys. Res. Lett.*, **42**, 6819–
34 6828, doi:10.1002/2015GL064924. <http://doi.wiley.com/10.1002/2015GL064924> (Accessed September
35 22, 2018).
- 36 Williams, J., and P. J. Crutzen, 2010: Nitrous oxide from aquaculture. *Nat. Geosci.*, **3**, 143,
37 doi:10.1038/ngeo804. <http://dx.doi.org/10.1038/ngeo804>.
- 38 Williams, J. W., and S. T. Jackson, 2007: Novel climates, no-analog communities, and ecological surprises.
39 *Front. Ecol. Environ.*, **5**, 475–482, doi:10.1890/070037. <http://doi.wiley.com/10.1890/070037>
40 (Accessed April 6, 2019).
- 41 Williams, J. W., S. T. Jackson, and J. E. Kutzbach, 2007: Projected distributions of novel and disappearing
42 climates by 2100 AD. *Proc. Natl. Acad. Sci.*, **104**, 5738–5742, doi:10.1073/pnas.0606292104.
43 <http://www.pnas.org/cgi/doi/10.1073/pnas.0606292104> (Accessed May 19, 2018).
- 44 Wilson, D., C. A. Farrell, D. Fallon, G. Moser, C. Müller, and F. Renou-Wilson, 2016: Multiyear greenhouse
45 gas balances at a rewetted temperate peatland. *Glob. Chang. Biol.*, **22**, 4080–4095,
46 doi:10.1111/gcb.13325.
- 47 Winsemius, H. C., and Coauthors, 2016: Global drivers of future river flood risk. *Nat. Clim. Chang.*, **6**, 381–
48 385, doi:10.1038/nclimate2893.
- 49 Wobus, C., and Coauthors, 2017: Climate change impacts on flood risk and asset damages within mapped
50 100-year floodplains of the contiguous United States. *Nat. Hazards Earth Syst. Sci.*, **17**, 2199–2211,
51 doi:10.5194/nhess-17-2199-2017. <https://www.nat-hazards-earth-syst-sci.net/17/2199/2017/> (Accessed
52 October 30, 2018).
- 53 Wolf, A., W. R. L. Anderegg, and S. W. Pacala, 2016a: Optimal stomatal behavior with competition for
54 water and risk of hydraulic impairment. *Proc. Natl. Acad. Sci.*, **113**, E7222–E7230,
55 doi:10.1073/pnas.1615144113. <http://www.pnas.org/lookup/doi/10.1073/pnas.1615144113>.
- 56 Wolf, J., G. R. Asrar, and T. O. West, 2017: Revised methane emissions factors and spatially distributed

- 1 annual carbon fluxes for global livestock. *Carbon Balance Manag.*, **12**, 16, doi:10.1186/s13021-017-
2 0084-y. <https://cbmjournals.biomedcentral.com/articles/10.1186/s13021-017-0084-y>.
- 3 Wolf, S., and Coauthors, 2016b: Warm spring reduced carbon cycle impact of the 2012 US summer drought.
4 *Proc. Natl. Acad. Sci.*, **113**, 5880–5885, doi:10.1073/pnas.1519620113.
5 <http://www.ncbi.nlm.nih.gov/pubmed/27114518> (Accessed October 30, 2018).
- 6 Woods, J., L. M. B. M. K. L. Kline, and A. Faaij, 2015: *Chapter 9, Land and bioenergy . Bioenergy &*
7 *sustainability: bridging the gaps*.
- 8 Woolf, D., J. E. Amonette, F. A. Street-Perrott, J. Lehmann, and S. Joseph, 2010: Sustainable biochar to
9 mitigate global climate change. *Nat. Commun.*, **1**, doi:10.1038/ncomms1053.
- 10 Worden, J. R., A. A. Bloom, S. Pandey, Z. Jiang, H. M. Worden, T. W. Walker, S. Houweling, and T.
11 Röckmann, 2017: Reduced biomass burning emissions reconcile conflicting estimates of the post-2006
12 atmospheric methane budget. *Nat. Commun.*, **8**, 1–12, doi:10.1038/s41467-017-02246-0.
- 13 Wright, S. J., and Coauthors, 2018: Plant responses to fertilization experiments in lowland, species-rich,
14 tropical forests. *Ecology*, **99**, 1129–1138.
- 15 Wu, M., G. Schurgers, M. Rummukainen, B. Smith, P. Samuelsson, C. Jansson, J. Siltberg, and W. May,
16 2016: Vegetation-climate feedbacks modulate rainfall patterns in Africa under future climate change.
17 *Earth Syst. Dyn.*, **7**, 627–647, doi:10.5194/esd-7-627-2016.
- 18 Wu, M., G. Schurgers, A. Ahlström, M. Rummukainen, P. A. Miller, B. Smith, and W. May, 2017: Impacts
19 of land use on climate and ecosystem productivity over the Amazon and the South American continent
20 Impacts of land use on climate and ecosystem productivity over the Amazon and the South American
21 continent. *Environ. Res. Lett.*, **12**.
- 22 Wu, P., N. Christidis, and P. Stott, 2013: Anthropogenic impact on Earth’s hydrological cycle. *Nat. Clim.*
23 *Chang.*, **3**, 807–810, doi:10.1038/nclimate1932.
24 <http://www.nature.com/doi/10.1038/nclimate1932> (Accessed May 16, 2018).
- 25 Wu, W., and Coauthors, 2019: Global advanced bioenergy potential under environmental protection policies
26 and societal transformation measures. *GCB Bioenergy*, doi:10.1111/gcbb.12614.
- 27 Wuebbles, D., and K. Hayhoe, 2002: Atmospheric methane and global change. *Earth-Science Rev.*, **57**, 177–
28 210, doi:10.1016/S0012-8252(01)00062-9.
- 29 Wypych, A., Z. Ustrnul, A. Sulikowska, F. M. Chmielewski, and B. Bochenek, 2017: Spatial and temporal
30 variability of the frost-free season in Central Europe and its circulation background. *Int. J. Climatol.*,
31 **37**, 3340–3352, doi:10.1002/joc.4920.
- 32 Xu, C., H. Liu, A. P. Williams, Y. Yin, and X. Wu, 2016a: Trends toward an earlier peak of the growing
33 season in Northern Hemisphere mid-latitudes. *Glob. Chang. Biol.*, **22**, 2852–2860,
34 doi:10.1111/gcb.13224. <http://doi.wiley.com/10.1111/gcb.13224> (Accessed January 23, 2018).
- 35 Xu, H. jie, and X. ping Wang, 2016: Effects of altered precipitation regimes on plant productivity in the arid
36 region of northern China. *Ecol. Inform.*, **31**, 137–146, doi:10.1016/j.ecoinf.2015.12.003.
37 <https://www.sciencedirect.com/science/article/pii/S1574954115001995> (Accessed November 2, 2018).
- 38 Xu, L., A. Samanta, M. H. Costa, S. Ganguly, R. R. Nemani, and R. B. Myneni, 2011: Widespread decline in
39 greenness of Amazonian vegetation due to the 2010 drought. *Geophys. Res. Lett.*, **38**, n/a-n/a,
40 doi:10.1029/2011GL046824. <http://doi.wiley.com/10.1029/2011GL046824> (Accessed May 7, 2018).
- 41 Xu, Q., R. Yang, Y.-X. X. Dong, Y.-X. X. Liu, and L.-R. R. Qiu, 2016b: The influence of rapid urbanization
42 and land use changes on terrestrial carbon sources/sinks in Guangzhou, China. *Ecol. Indic.*, **70**, 304–
43 316, doi:10.1016/j.ecolind.2016.05.052.
44 <http://www.sciencedirect.com/science/article/pii/S1470160X16303120>.
- 45 Xu, X., and Coauthors, 2016c: Four decades of modeling methane cycling in terrestrial ecosystems.
46 *Biogeosciences*, **13**, 3735–3755, doi:10.5194/bg-13-3735-2016.
47 <https://www.biogeosciences.net/13/3735/2016/>.
- 48 Xu, X., W. J. Riley, C. D. Koven, and G. Jia, 2018: Observed and Simulated Sensitivities of Spring Greenup
49 to Preseason Climate in Northern Temperate and Boreal Regions. *J. Geophys. Res. Biogeosciences*,
50 **123**, 60–78, doi:10.1002/2017JG004117. <http://doi.wiley.com/10.1002/2017JG004117>.
- 51 Xue, Y., and Coauthors, 2016: West African monsoon decadal variability and surface-related forcings:
52 second West African Monsoon Modeling and Evaluation Project Experiment (WAMME II). *Clim.*
53 *Dyn.*, **47**, 3517–3545, doi:10.1007/s00382-016-3224-2.
- 54 Yamori, W., K. Hikosaka, and D. A. Way, 2014: Temperature response of photosynthesis in C3, C4, and
55 CAM plants: Temperature acclimation and temperature adaptation. *Photosynth. Res.*, **119**, 101–117,
56 doi:10.1007/s11120-013-9874-6.

- 1 Yan, M., J. Liu, and Z. Wang, 2017: Global climate responses to land use and land cover changes over the
2 past two millennia. *Atmosphere (Basel)*, doi:10.3390/atmos8040064.
- 3 Yan, Z., B. Bond-Lamberty, K. E. Todd-Brown, V. L. Bailey, S. Li, C. Liu, and C. Liu, 2018: A moisture
4 function of soil heterotrophic respiration that incorporates microscale processes. *Nat. Commun.*, **9**,
5 2562, doi:10.1038/s41467-018-04971-6.
- 6 Yang, H., Q. Li, K. Wang, Y. Sun, and D. Sun, 2015a: Decomposing the meridional heat transport in the
7 climate system. *Clim. Dyn.*, **44**, 2751–2768, doi:10.1007/s00382-014-2380-5.
8 <http://link.springer.com/10.1007/s00382-014-2380-5> (Accessed May 7, 2018).
- 9 Yang, L., G. Sun, L. Zhi, and J. Zhao, 2018: Negative soil moisture-precipitation feedback in dry and wet
10 regions. *Sci. Rep.*, **8**, 4026, doi:10.1038/s41598-018-22394-7.
- 11 Yang, X., and Coauthors, 2015b: Potential benefits of climate change for crop productivity in China. *Agric.*
12 *For. Meteorol.*, **208**, 76–84, doi:10.1016/j.agrformet.2015.04.024. [https://www-sciencedirect-](https://www-sciencedirect-com.ezproxy.uct.ac.za/science/article/pii/S0168192315001288)
13 [com.ezproxy.uct.ac.za/science/article/pii/S0168192315001288](https://www-sciencedirect-com.ezproxy.uct.ac.za/science/article/pii/S0168192315001288) (Accessed May 10, 2018).
- 14 Yao, Y., S. Piao, and T. Wang, 2018: Future biomass carbon sequestration capacity of Chinese forests. *Sci.*
15 *Bull.*, doi:10.1016/j.scib.2018.07.015.
- 16 Ye, J. S., Y. H. Gong, F. Zhang, J. Ren, X. K. Bai, and Y. Zheng, 2018: Which temperature and precipitation
17 extremes best explain the variation of warm versus cold years and wet versus dry years? *J. Clim.*, **31**,
18 45–59, doi:10.1175/JCLI-D-17-0377.1. <http://journals.ametsoc.org/doi/10.1175/JCLI-D-17-0377.1>
19 (Accessed May 17, 2018).
- 20 Yellen, B., J. D. Woodruff, L. N. Kratz, S. B. Mabee, J. Morrison, and A. M. Martini, 2014: Source,
21 conveyance and fate of suspended sediments following Hurricane Irene. New England, USA.
22 *Geomorphology*, **226**, 124–134, doi:10.1016/j.geomorph.2014.07.028. [https://www-sciencedirect-](https://www-sciencedirect-com.ezproxy.uct.ac.za/science/article/pii/S0169555X14003936)
23 [com.ezproxy.uct.ac.za/science/article/pii/S0169555X14003936](https://www-sciencedirect-com.ezproxy.uct.ac.za/science/article/pii/S0169555X14003936) (Accessed May 25, 2018).
- 24 Yim, S. Y., B. Wang, J. Liu, and Z. Wu, 2014: A comparison of regional monsoon variability using monsoon
25 indices. *Clim. Dyn.*, **43**, 1423–1437, doi:10.1007/s00382-013-1956-9.
- 26 Yin, Y., Q. Tang, L. Wang, and X. Liu, 2016: Risk and contributing factors of ecosystem shifts over
27 naturally vegetated land under climate change in China. *Sci. Rep.*, **6**, 1–11, doi:10.1038/srep20905.
28 <http://dx.doi.org/10.1038/srep20905>.
- 29 Yin, Y., D. Ma, and S. Wu, 2019: Enlargement of the semi-arid region in China from 1961 to 2010. *Clim.*
30 *Dyn.*, **52**, 509–521, doi:10.1007/s00382-018-4139-x. [http://link.springer.com/10.1007/s00382-018-](http://link.springer.com/10.1007/s00382-018-4139-x)
31 [4139-x](http://link.springer.com/10.1007/s00382-018-4139-x) (Accessed April 15, 2019).
- 32 York, R., 2012: Do alternative energy sources displace fossil fuels? *Nat. Clim. Chang.*,
33 doi:10.1038/nclimate1451.
- 34 Yu, L., Y. Huang, W. Zhang, T. Li, and W. Sun, 2017: Methane uptake in global forest and grassland soils
35 from 1981 to 2010. *Sci. Total Environ.*, **607–608**, 1163–1172, doi:10.1016/j.scitotenv.2017.07.082.
36 <http://dx.doi.org/10.1016/j.scitotenv.2017.07.082>.
- 37 Yu, M., G. Wang, and H. Chen, 2016: Quantifying the impacts of land surface schemes and dynamic
38 vegetation on the model dependency of projected changes in surface energy and water budgets. *J. Adv.*
39 *Model. Earth Syst.*, **8**, 370–386, doi:10.1002/2015MS000492. <https://doi.org/10.1002/2015MS000492>.
- 40 Yuan, X., L. Wang, and E. . Wood, 2018: Anthropogenic intensification of southern African flash droughts
41 as exemplified by the 2015/16 season. *Bull. Am. Meteorol. Soc.*, **99**, S86–S90.
- 42 Yue, X., and N. Unger, 2018: Fire air pollution reduces global terrestrial productivity. *Nat. Commun.*, **9**,
43 5413, doi:10.1038/s41467-018-07921-4. <https://doi.org/10.1038/s41467-018-07921-4>.
- 44 Zaitchik, B. F., A. K. Macalady, L. R. Bonneau, and R. B. Smith, 2006: Europe’s 2003 heat wave: A satellite
45 view of impacts and land - Atmosphere feedbacks. *Int. J. Climatol.*, **26**, 743–769,
46 doi:10.1002/joc.1280. <https://doi.org/10.1002/joc.1280>.
- 47 Zak, D. R., K. S. Pregitzer, M. E. Kubiske, and A. J. Burton, 2011: Forest productivity under elevated CO₂
48 and O₃: Positive feedbacks to soil N cycling sustain decade-long net primary productivity enhancement
49 by CO₂. *Ecol. Lett.*, **14**, 1220–1226, doi:10.1111/j.1461-0248.2011.01692.x.
50 <https://doi.org/10.1111/j.1461-0248.2011.01692.x>.
- 51 Zampieri, M., A. Ceglar, F. Dentener, and A. Toretì, 2017: Wheat yield loss attributable to heat waves,
52 drought and water excess at the global, national and subnational scales. *Environ. Res. Lett.*, **12**, 64008,
53 doi:10.1088/1748-9326/aa723b.
- 54 Zarin, D. J., and Coauthors, 2016a: Can carbon emissions from tropical deforestation drop by 50% in
55 5 years? *Glob. Chang. Biol.*, **22**, 1336–1347, doi:10.1111/gcb.13153.
- 56 —, and Coauthors, 2016b: Can carbon emissions from tropical deforestation drop by 50% in 5 years?

- 1 *Glob. Chang. Biol.*, **22**, 1336–1347, doi:10.1111/gcb.13153. <https://doi.org/10.1111/gcb.13153>.
- 2 Zelenáková, M., D. C. Diaconu, and K. Haarstad, 2017: Urban Water Retention Measures. *Procedia Eng.*,
3 **190**, 419–426, doi:10.1016/j.proeng.2017.05.358.
- 4 Zemp, D. C., C. Schleussner, H. M. J. Barbosa, R. J. Van Der Ent, J. F. Donges, J. Heinke, G. Sampaio, and
5 C. Analytics, 2014: On the importance of cascading moisture recycling in South America. 13337–
6 13359, doi:10.5194/acp-14-13337-2014.
- 7 Zeng, Z., and Coauthors, 2017: Climate mitigation from vegetation biophysical feedbacks during the past
8 three decades. *Nat. Clim. Chang.*, **7**, 432–436, doi:10.1038/nclimate3299.
- 9 Zeppel, M. J. B., J. V. Wilks, and J. D. Lewis, 2014: Impacts of extreme precipitation and seasonal changes
10 in precipitation on plants. *Biogeosciences*, **11**, 3083–3093, doi:10.5194/bg-11-3083-2014.
11 <https://www.biogeosciences.net/11/3083/2014/> (Accessed November 2, 2018).
- 12 Zhang, D., H. Zheng, Z. Ren, C. Zhai, G. Shen, Z. Mao, P. Wang, and X. He, 2015: Effects of forest type
13 and urbanization on carbon storage of urban forests in Changchun, Northeast China. *Chinese Geogr.*
14 *Sci.*, **25**, 147–158, doi:10.1007/s11769-015-0743-4.
- 15 Zhang, G., and Coauthors, 2017: Extensive and drastically different alpine lake changes on Asia's high
16 plateaus during the past four decades. *Geophys. Res. Lett.*, **44**, 252–260, doi:10.1002/2016GL072033.
- 17 Zhang, H. L., R. Lal, X. Zhao, J. F. Xue, and F. Chen, 2014a: Opportunities and challenges of soil carbon
18 sequestration by conservation agriculture in China. *Advances in Agronomy*, Vol. 124 of, 1–36.
- 19 Zhang, M., and Coauthors, 2014b: Response of surface air temperature to small-scale land clearing across
20 latitudes. *Environ. Res. Lett.*, doi:10.1088/1748-9326/9/3/034002.
- 21 Zhang, Q., A. J. Pitman, Y. P. Wang, Y. J. Dai, and P. J. Lawrence, 2013a: The impact of nitrogen and
22 phosphorous limitation on the estimated terrestrial carbon balance and warming of land use change
23 over the last 156 yr. *Earth Syst. Dyn.*, **4**, 333–345, doi:10.5194/esd-4-333-2013.
- 24 Zhang, W. -f., and Coauthors, 2013b: New technologies reduce greenhouse gas emissions from nitrogenous
25 fertilizer in China. *Proc. Natl. Acad. Sci.*, **110**, 8375–8380, doi:10.1073/pnas.1210447110.
- 26 ———, and Coauthors, 2013c: New technologies reduce greenhouse gas emissions from nitrogenous fertilizer
27 in China. *Proc. Natl. Acad. Sci.*, **110**, 8375–8380, doi:10.1073/pnas.1210447110.
28 <http://www.pnas.org/cgi/doi/10.1073/pnas.1210447110>.
- 29 Zhang, Y., and Coauthors, 2019: Increased Global Land Carbon Sink Due to Aerosol-Induced Cooling.
30 *Global Biogeochem. Cycles*, **0**, doi:10.1029/2018GB006051.
- 31 Zhao, C., and Coauthors, 2017: Temperature increase reduces global yields of major crops in four
32 independent estimates. *Proc. Natl. Acad. Sci.*, 201701762, doi:10.1073/pnas.1701762114.
- 33 Zhao, G., B. A. Bryan, D. King, Z. Luo, E. Wang, and Q. Yu, 2015: Sustainable limits to crop residue
34 harvest for bioenergy: Maintaining soil carbon in Australia's agricultural lands. *GCB Bioenergy*,
35 doi:10.1111/gcbb.12145.
- 36 Zhao, K., and R. B. Jackson, 2014: Biophysical forcings of land-use changes from potential forestry
37 activities in North America. *Ecol. Monogr.*, **84**, 329–353, doi:10.1890/12-1705.1.
- 38 Zhao, L., X. Lee, R. B. Smith, and K. Oleson, 2014: Strong contributions of local background climate to
39 urban heat islands. *Nature*, **511**, 216–219, doi:10.1038/nature13462.
- 40 Zhao, L., A. Dai, and B. Dong, 2018: Changes in global vegetation activity and its driving factors during
41 1982–2013. *Agric. For. Meteorol.*, **249**, 198–209, doi:10.1016/j.agrformet.2017.11.013.
42 <https://www.sciencedirect.com/science/article/pii/S0168192317303866> (Accessed February 4, 2018).
- 43 Zhao, M., and A. J. Pitman, 2002: The impact of land cover change and increasing carbon dioxide on the
44 extreme and frequency of maximum temperature and convective precipitation. *Geophys. Res. Lett.*,
45 doi:10.1029/2001gl013476.
- 46 Zhong, S., and Coauthors, 2017: Urbanization-induced urban heat island and aerosol effects on climate
47 extremes in the Yangtze River Delta region of China. *Atmos. Chem. Phys.*, **17**, 5439–5457,
48 doi:10.5194/acp-17-5439-2017.
- 49 Zhou, B., D. Rybski, and J. P. Kropp, 2017: The role of city size and urban form in the surface urban heat
50 island. *Sci. Rep.*, **7**, 4791, doi:10.1038/s41598-017-04242-2. <https://doi.org/10.1038/s41598-017-04242-2>.
- 51
- 52 Zhou, D., L. Zhang, L. Hao, G. Sun, Y. Liu, and C. Zhu, 2016: Spatiotemporal trends of urban heat island
53 effect along the urban development intensity gradient in China. *Sci. Total Environ.*, **544**, 617–626,
54 doi:10.1016/j.scitotenv.2015.11.168. <http://dx.doi.org/10.1016/j.scitotenv.2015.11.168>.
- 55 Zhou, J., and Coauthors, 2011: Microbial mediation of carbon-cycling feedbacks to climate warming. *Nat.*
56 *climate Chang.*, **2**, 106–110. <http://dx.doi.org/10.1038/nclimate1331>.

- 1 Zhou, L., and Coauthors, 2014: Widespread decline of Congo rainforest greenness in the past decade.
2 *Nature*, **508**, 86–90, doi:10.1038/nature13265.
- 3 Zhou, Y., and Coauthors, 2015: A global map of urban extent from nightlights. *Environ. Res. Lett.*, **10**,
4 2000–2010, doi:10.1088/1748-9326/10/5/054011.
- 5 Zhu, J., M. Zhang, Y. Zhang, X. Zeng, and X. Xiao, 2018: Response of tropical terrestrial gross primary
6 production to the super El Niño event in 2015. *J. Geophys. Res. Biogeosciences*,
7 doi:10.1029/2018JG004571.
- 8 Zhu, P., Q. Zhuang, P. Ciais, L. Welp, W. Li, and Q. Xin, 2017: Elevated atmospheric CO₂ negatively
9 impacts photosynthesis through radiative forcing and physiology-mediated climate feedback. *Geophys.*
10 *Res. Lett.*, **44**, 1956–1963, doi:10.1002/2016GL071733.
- 11 Zhu, Z., and Coauthors, 2016: Greening of the Earth and its drivers. *Nat. Clim. Chang.*, **6**, 791–795,
12 doi:10.1038/nclimate3004.
- 13 Zhuang, Q., and Coauthors, 2006: CO₂ and CH₄ exchanges between land
14 ecosystems and the atmosphere in northern high latitudes over the 21st century. *Geophys. Res. Lett.*, **33**,
15 2–6, doi:10.1029/2006GL026972.
- 16 ———, and Coauthors, 2013: Response of global soil consumption of atmospheric methane to changes in
17 atmospheric climate and nitrogen deposition. *Global Biogeochem. Cycles*, **27**, 650–663,
18 doi:10.1002/gbc.20057.
- 19 Zilberman, D., 2017: Indirect land use change: much ado about (almost) nothing. *GCB Bioenergy*, **9**, 485–
20 488, doi:10.1111/gcbb.12368. <http://dx.doi.org/10.1111/gcbb.12368>.
- 21 Zilli, M. T., L. M. V. Carvalho, B. Liebmann, and M. A. Silva Dias, 2017: A comprehensive analysis of
22 trends in extreme precipitation over southeastern coast of Brazil. *Int. J. Climatol.*, **37**, 2269–2279,
23 doi:10.1002/joc.4840.
- 24 Zipper, S. C., J. Schatz, C. J. Kucharik, and S. P. Loheide, 2017: Urban heat island-induced increases in
25 evapotranspirative demand. *Geophys. Res. Lett.*, **44**, 873–881, doi:10.1002/2016GL072190.
- 26 Zölch, T., J. Maderspacher, C. Wamsler, and S. Pauleit, 2016: Using green infrastructure for urban climate-
27 proofing: An evaluation of heat mitigation measures at the micro-scale. *Urban For. Urban Green.*, **20**,
28 305–316, doi:<https://doi.org/10.1016/j.ufug.2016.09.011>.
29 <http://www.sciencedirect.com/science/article/pii/S1618866716301686>.
- 30 Zomer, R. J., H. Neufeldt, J. Xu, A. Ahrends, D. Bossio, A. Trabucco, M. van Noordwijk, and M. Wang,
31 2016: Global Tree Cover and Biomass Carbon on Agricultural Land: The contribution of agroforestry
32 to global and national carbon budgets. *Sci. Rep.*, **6**, 29987. <https://doi.org/10.1038/srep29987>.
- 33 Zomer, R. J., D. A. Bossio, R. Sommer, and L. V. Verchot, 2017: Global Sequestration Potential of
34 Increased Organic Carbon in Cropland Soils. *Sci. Rep.*, doi:10.1038/s41598-017-15794-8.
- 35 Zscheischler, J., and Coauthors, 2018: Future climate risk from compound events. *Nat. Clim. Chang.*, **1**,
36 doi:10.1038/s41558-018-0156-3.
- 37
38
39
40
41

Appendix

This appendix provides all numbers that support **Figure 2.14** and **Figure 2.17** located in Section 2.5. It lists all model-based studies, with their references, that have been used to create the figures. Studies that examine the effects of historical and future scenarios of changes in anthropogenic land cover are presented in Table 1. The responses to idealised latitudinal deforestation and forestation can be found in Table A2.2.

The biophysical effects of changes in anthropogenic land cover reflect the impacts of changes in physical land surface characteristics such as albedo, evapotranspiration, and roughness length. The biogeochemical effects reflect changes in atmospheric CO₂ composition resulting from anthropogenic changes in land cover. The biogeochemical effects are estimated using three different methods:

1. Directly calculated within global climate models (Tables A2.1 and A2.2);
2. Calculated from off-line dynamic global vegetation models (DGVMs) estimates of net changes in the emissions of CO₂ from land (Table A2.1);
3. Calculated from observation-based estimates of net changes in the emissions of CO₂ from land (for historical reconstruction only, Table A2.1)

The mean annual and global temperature change (ΔT) resulting from biogeochemical effects is calculated as follows, for both DGVMs and observation-based estimates:

$$\Delta T = \Delta LCO_2 * TCRE$$

Where ΔLCO_2 is the cumulative changes in net emissions of CO₂ resulting from anthropogenic land cover changes during the time period considered (in Tera tons of carbon, TtC), and TCRE is the transient climate response to cumulative carbon emissions (Gillett et al. 2013; Matthews et al. 2009). TCRE is a measure of the global temperature response to cumulative emissions of CO₂ and has been identified as a useful and practical tool for evaluating CO₂-induced climate changes (expressed in °C per Tera tons of Carbon, °C/TtC). TCRE values have been estimated for a range of Earth system models (Gillett et al. 2013), (MacDougall et al. 2016). In the following, we use the 5th percentile, mean, and 95th percentile derived from the range of available TCRE values. For each DGVM or observation-based estimate, we then calculate three potential temperature change to bracket the range of climate sensitivities.

1
2
3
4
5
6
7

Table A2.1: Model-based and observation-based estimates of the effects historical and future anthropogenic land cover changes have on mean annual global surface air temperature (°C). BGC and BPH correspond to the change in temperature resulting from respectively biogeochemical processes (e.g., changes in atmospheric CO₂ composition) and biophysical processes (e.g., changes in physical land surface characteristics such as albedo, evapotranspiration, and roughness length).

| Reference of the study | Time period | Cumulative CO ₂ emissions from anthropogenic land cover change (TtC) | TCRE (°C/TtC) | Change in mean global annual (°C) | |
|--|---------------------------------------|---|---------------|-----------------------------------|---|
| | | | | BGC | BPH |
| <i>Historical period (global climate models)</i> | | | | | |
| (Lawrence et al. 2018) | 1850-2005 | 0.123 | 1.9 | 0.23 | |
| (Simmons and Matthews 2016) | 1750-2000 ¹⁰ | | | 0.22 | -0.24 |
| (Devaraju et al. 2016) | 1850-2005 | 0.112 | 1.9 | 0.21 | |
| (Zhang et al. 2013a) | 1850-2005 ¹¹ | 0.097 | 1.75 | 0.17 | -0.06 |
| (Hua and Chen 2013) | ~1850-2000 (average of two estimates) | | | | -0.015 |
| (Jones et al. 2013a) | Preindustrial (no exact dates) | | | | -0.57 |
| (Lawrence et al. 2012) | 1850-2005 | 0.120 | 1.9 | 0.23 | -0.10 |
| (De Noblet-Ducoudré et al. 2012) | 1972-2002 relative to 1900-1970 | | | | -0.042 ; -0.056 ; -0.005 ; -0.041 ; 0.021 ; -0.007 ; -0.005 |
| (Pongratz et al. 2010) | 20th century | | | 0.16, 0.18 | -0.03 |
| (Arora and Boer 2010) | 1850-2000 | 0.040,0.077 | 2.4 | 0.1, 0.18 | |
| (Strengers et al. 2010) | 20 th century | | | | -0.06 |
| (Kvalevåg et al. 2010) | Preindustrial (no exact dates) | | | | +0.04 (CASE I) |

¹⁰ FOOTNOTE: Land Use Change + Fossil Fuel emission simulation values are considered.

¹¹ FOOTNOTE: Carbon-Nitrogen-Phosphorous simulation values are considered.

| | | | | | |
|--|---|-------------|----------------|--------------------------------|---------------------------------------|
| (Findell et al. 2009) | 1901-2004 | | | | +0.02 |
| (Findell et al. 2007) | 1990 relative to potential vegetation | | | | +0.008 |
| (Brovkin et al. 2006) | 1700-1992 (5 models) | | | | -0.24 ; -0.13 ; -0.14 ; -0.25 ; -0.17 |
| (Betts et al. 2007; Betts 2001) | 1750-1990 | | | | -0.02 |
| (Hansen et al. 2005) | 1880-1990 | | | | -0.04 |
| (Feddema et al. 2005) | Preindustrial land-cover changes (no exact dates, "prehuman" simulations) | | | | -0.39 |
| (Matthews et al. 2004) | 1700-2000 (average of 7 simulations) | | | 0.3 | -0.14 |
| (Brovkin et al. 2004) | 1800-2000 | | | 0.18 | -0.26 |
| (Zhao and Pitman 2002; Chase et al. 2000, 2001) | Preindustrial | | | | +0.06 |
| (Hansen et al. 1998) | Preindustrial land-cover changes | | | | -0.14 |
| Mean (\pmStandard deviation) of all studies | | | | 0.2\pm0.05 | -0.1\pm0.14 |
| <i>Historical period (Dynamic Global Vegetation Models/Bookkeeping model results)</i> | | | | | |
| (Li et al. 2017a) | 1901-2012 (median of models) | 0.148 | 0.88-1.72-2.52 | 0.13-0.25-0.37 | |
| (Peng et al. 2017) | 1850-1990 (realistic cases range) | 0.087,0.139 | 0.88-1.72-2.52 | 0.1-0.15-0.22, 0.12-0.24-0.35 | |
| (Arneeth et al. 2017a) | 1901-2014 ¹² | 0.089 | 0.88-1.72-2.52 | 0.1-0.15-0.22 | |
| | | 0.210 | 0.88-1.72-2.52 | 0.18-0.36-0.53 | |
| | | 0.179 | 0.88- | 0.16-0.31- | |

¹² FOOTNOTE: FLULCC,1 refers to land use change related fluxes accounting for new processes in their study.

| | | | | | |
|--|--|--------------|----------------|---------------------------------|--|
| | | | 1.72-2.52 | 0.45 | |
| | | 0.195 | 0.88-1.72-2.52 | 0.17-0.33-0.49 | |
| | | 0.083 | 0.88-1.72-2.52 | 0.1-0.14-0.21 | |
| | | 0.161 | 0.88-1.72-2.52 | 0.14-0.28-0.4 | |
| | | 0.117 | 0.88-1.72-2.52 | 0.1-0.2-0.3 | |
| | | 0.104 | 0.88-1.72-2.52 | 0.1-0.18-0.26 | |
| | | 0.196 | 0.88-1.72-2.52 | 0.17-0.34-0.49 | |
| (Pugh et al. 2015) | 1850-2012 (gross land clearance flux) | 0.157 | 0.88-1.72-2.52 | 0.14-0.27-0.39 | |
| (Hansis et al. 2015) | 1850-2012 | 0.269 | 0.88-1.72-2.52 | 0.19-0.36-0.53 | |
| (Houghton et al. 2012a; Hansis et al. 2015) | 1920-1999 (multi-model range) | 0.072, 0.115 | 0.88-1.72-2.52 | 0.1-0.12-0.18, 0.1-0.2-0.3 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.24\pm0.12 | |
| <i>Historical period (Observation-based estimates)</i> | | | | | |
| (Li et al. 2017a) | 1901-2012 | 0.155 | 0.88-1.72-2.52 | 0.14-0.27-0.39 | |
| (Li et al. 2017a; Avitabile et al. 2016; Carvalhais et al. 2014) | 1901-2012 ¹³ | 0.160,0.165 | 0.88-1.72-2.52 | 0.14-0.27-0.40,0.14-0.28-0.41 | |
| (Liu et al. 2015; Li et al. 2017a) | 1901-2012 | 0.161,0.163 | 0.88-1.72-2.52 | 0.14-0.28-0.41 | |
| (Le Quéré et al. 2015) | 1870-2014 | 0.145 | 0.88-1.72-2.52 | 0.13-0.25-0.36 | |
| (Carvalhais et al. 2014; Li et al. 2017a) | 1901-2012 | 0.152,0.159 | 0.88-1.72-2.52 | 0.13-0.26-0.38, 0.14-0.27-0.4 | |
| (Pan et al. 2011b; Li et al. 2017a) | 1901-2012 | 0.119,0.122 | 0.88-1.72-2.52 | 0.10-0.20-0.30, 0.11- | |

¹³ FOOTNOTE: Different harmonization methods. Method A assumes increase in cropland area in a grid cell taken from forest. Method C assumes increase in cropland and pasture taken from forest and then natural grassland if no more forest area available.

| | | | | | |
|--|-----------|-------------|----------------|---------------------------------|---------------------------------|
| | | | | 0.21-0.31 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.25\pm0.10 | |
| Future -RCP8.5 (global climate models) | | | | | |
| (Tharammal et al. 2018) | 2006-2100 | 0.093 | 1.9 | 0.18 | |
| (Lawrence et al. 2018) | 2006-2100 | 0.211 | 1.9 | 0.40 | |
| (Simmons and Matthews 2016) | 2000-2100 | - | - | 0.35 | -0.34 |
| (Hua et al. 2015) | 2006-2100 | 0.032 | 2.4 | 0.08 | - |
| (Davies-Barnard et al. 2014) | 2005-2100 | 0.02 | 2.1 | 0.04 | -0.015 |
| (Boysen et al. 2014; Quesada et al. 2017a; Brovkin et al. 2013b) | 2005-2100 | 0.034 | 2.4 | 0.08 | 0.04 |
| | | 0.025 | 2.1 | 0.05 | 0.0 |
| | | 0.037 | 1.6 | 0.06 | 0.08 |
| | | 0.062 | 2.2 | 0.13 | -0.20 |
| | | 0.205 | 1.6 | 0.33 | -0.06 |
| (Lawrence et al. 2012) | 2006-2100 | 0.256 | 1.9 | 0.49 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.20\pm0.15 | -0.1\pm0.14 |
| Future -RCP8.5 (Dynamic Global Vegetation Model results) | | | | | |
| (Pugh et al. 2015) | 2006-2100 | 0.169,0.171 | 0.88-1.72-2.52 | 0.15-0.29-0.42,0.15-0.29-0.43 | |
| (IPCC 2014) | 2005-2099 | 0.151 | 0.88-1.72-2.52 | 0.13-0.26-0.38 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.28\pm0.11 | |
| Future RCP4.5 (global climate models) | | | | | |
| (Tharammal et al. 2018) | 2005-2100 | -0.029 | 1.9 | -0.05 | |
| (Lawrence et al. 2018) | 2006-2100 | 0.053 | 1.9 | 0.10 | |
| (Simmons and Matthews 2016) | 2000-2100 | | | 0.37 | -0.29 |
| (Davies-Barnard et al. 2014) | 2005-2100 | -0.040 | 2.1 | -0.08 | 0.14 |

| | | | | | |
|---|---------------------------------------|--------------|----------------|---------------------------------------|----------------------------------|
| (Lawrence et al. 2012) | 2006-2100 | 0.148 | 1.9 | 0.28 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.12\pm0.17 | -0.1\pm0.21 |
| Future RCP4.5 (Dynamic Global Vegetation Model results) | | | | | |
| (Pugh et al. 2015) | 2006-2100 | 0.016,-0.018 | 0.88-1.72-2.52 | 0.01-0.03-0.04,-0.02-(-0.03)-(-0.045) | |
| (IPCC 2014) | 2005-2099 | 0.027 | 0.88-1.72-2.52 | 0.02-0.05-0.07 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.01\pm0.04 | |
| Future RCP2.6 (global climate models) | | | | | |
| (Tharammal et al. 2018) | 2005-2100 | 0.039 | 1.9 | 0.07 | |
| (Simmons and Matthews 2016) | 2000-2100 | | | 0.42 | -0.35 |
| (Hua et al. 2015) | 2006-2100 | 0.036 | 2.4 | 0.09 | |
| (Davies-Barnard et al. 2014) | 2005-2100 | | | 0.04 | -0.01 |
| (Brovkin et al. 2013b) | 2005-2100 | 0.039 | 2.4 | 0.09 | |
| | | 0.019 | 2.1 | 0.04 | |
| | | 0.065 | 2.2 | 0.14 | |
| | | 0.175 | 1.6 | 0.28 | |
| (Lawrence et al. 2012) | 2006-2100 | 0.0154 | 1.9 | 0.03 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.13\pm0.12 | -0.18\pm0.17 |
| Future RCP2.6 (Dynamic Global Vegetation Model results) | | | | | |
| (Pugh et al. 2015) | 2006-2100 (no harvest, managed cases) | 0.057,0.084 | 0.88-1.72-2.52 | 0.05-0.09-0.14,0.07-0.14-0.21 | |
| (IPCC 2014) | 2005-2099 | 0.105 | 0.88-1.72-2.52 | 0.09-0.18-0.26 | |
| Mean (\pmStandard deviation) of all studies | | | | 0.14\pm0.06 | |

1

1 **Table A2.2: Model-based estimates of the effects idealised and latitudinal deforestation or forestation**
 2 **have on mean annual global and latitudinal surface air temperature (°C). BGC and BPH correspond**
 3 **to the change in temperature resulting from respectively biogeochemical processes (e.g. changes in**
 4 **atmospheric CO2 composition) and biophysical processes (e.g. changes in physical land surface**
 5 **characteristics such as albedo, evapotranspiration, and roughness length).**

| Idealised Deforestation/Afforestation (global climate models) | | | | | |
|---|---|---------------------------|--------------|---|-------------------|
| Reference | Change in forest area (Mkm ²) | Cumulative LCC flux (TtC) | TCRE (K/TtC) | Mean annual change in surface air temperature, averaged globally (and for the latitudinal band where trees are removed or added) (°C) | |
| | | | | BGC | BPH |
| <i>Tropical Deforestation</i> | | | | | |
| (Devaraju et al. 2018b) | 36.1 | | | | 0.02 (1.14) |
| (Longobardi et al. 2016b) | 23 ¹⁴ | 0.127 | 1.72 | 0.30 | 0.044 (-0.19) |
| (Devaraju et al. 2015c) | 23 | | | 1.06 | -0.04 (0.20) |
| (Brovkin et al. 2015) | | | | | -0.01,-0.13,-0.05 |
| (Bathiany et al. 2010b) | 23.1 | | | 0.40 | 0.18 (0.9) |
| (Snyder 2010) | 23 | | | | 0.2 (1.0) |
| (Bala et al. 2007) | 23 | 0.418 | 1.72 | 0.72 | 0.70 |
| (Voldoire 2006) | | | | | 0.2,0.4,0.6 |
| (Snyder et al. 2004b) | 22.7 | | | | 0.24(1.2) |
| (Claussen et al. 2001b) | 7.5 | | | 0.19 (0.15) | -0.04 (0.13) |
| (Ganopolski et al. 2001c) | 7.5 | | | | -0.5 (0.5) |
| (Henderson-Sellers and Gornitz 1984) | | | | | 0.00 |
| (Potter et al. 1981; POTTER et al. 1975) | | | | | -0.2 |
| (Sagan et al. | | | | | -0.07 |

¹⁴ FOOTNOTE: For some studies that do not provide area deforested, IPSL-CM5 model grids used to calculate the area.

| | | | | | |
|---|------|-------|------|-----------------|------------------------------------|
| 1979) | | | | | |
| Mean (\pm Standard deviation) of all studies | | | | 0.53 \pm 0.32 | 0.1 \pm 0.27(0.61 \pm 0.48) |
| Tropical Afforestation | | | | | |
| (Wang et al. 2014a) (Average of 4 simulations) | | | | | 0.925 |
| (Bathiany et al. 2010b) | 23.1 | | | | -0.03 (-0.1) |
| Temperate Deforestation | | | | | |
| (Devaraju et al. 2018a) | 18.8 | | | | 0.18 (0.52) |
| (Longobardi et al. 2016b) | 15 | 0.047 | 1.72 | 0.10 | -0.077 (-0.22) |
| (Devaraju et al. 2015b) | 15.3 | | | 0.39 | -0.5 (-0.8) |
| (Bala et al. 2007) | 15 | 0.231 | 1.72 | 0.40 | -0.04 |
| (Snyder et al. 2004b) | 19.1 | | | | -0.22 (-1.1) |
| Mean (\pm Standard deviation) of all studies | | | | 0.29 \pm 0.13 | -0.13 \pm 0.22 (-0.4 \pm 0.62) |
| Temperate Afforestation | | | | | |
| (Laguë and Swann 2016b) | | | | | 0.3 (1.5) |
| (Wang et al. 2014a) | | | | | 1.14 |
| (Swann et al. 2012b) | 15.3 | | | -0.2, -0.7 | 0.3 |
| (Gibbard et al. 2005) | | | | | 0.27 |
| Mean (\pm Standard deviation) of all studies | | | | -0.45 | 0.50 \pm 0.36 |
| Boreal Deforestation | | | | | |
| (Devaraju et al. 2018a) | 23.5 | | | | -0.25 (-1.2) |
| (Longobardi et al. 2016b) | 13.7 | 0.050 | 1.72 | 0.11 | -0.38 (-0.9) |

| | | | | | |
|---|------|--------|------|-----------------|-----------------------------------|
| (Devaraju et al. 2015b) | 13.7 | | | 0.06 | -0.9 (-4) |
| (Dass et al. 2013) | 18.5 | | | 0.12, 0.32 | -0.35 |
| (Bathiany et al. 2010b) | 18.5 | 0.02 | 2.04 | 0.04 | -0.28 (-1.1) |
| (Bala et al. 2007) | 13.7 | 0.0105 | 1.72 | 0.02 | -0.8 |
| (Snyder et al. 2004b) | 22.4 | | | | -0.77(-2.8) |
| (Caussen et al. 2001) | 6 | | | 0.09 (0.12) | -0.23 (-0.82) |
| (Ganopolski et al. 2001c) | 6 | | | | -1.0 |
| Mean (\pm Standard deviation) of all studies | | | | 0.11 \pm 0.09 | -0.55 \pm 0.29 (-1.8 \pm 1.2) |
| Boreal Afforestation | | | | | |
| (Bathiany et al. 2010b) | | | | | 0.31 (1.2) |

1
2

Chapter 3 : Desertification

1
2
3
4
5
6
7
8
9
10
11
12
13
14
15
16
17
18
19
20

Coordinating Lead Authors: Alisher Mirzabaev (Germany/Uzbekistan), Jianguo Wu (China)

Lead Authors: Jason Evans (Australia), Felipe Garcia-Oliva (Mexico), Ismail Abdel Galil Hussein (Egypt), Muhammad Mohsin Iqbal (Pakistan), Joyce Kimutai (Kenya), Tony Knowles (South Africa), Francisco Meza (Chile), Dalila Nedjraoui (Algeria), Fasil Tena (Ethiopia), Murat Türkeş (Turkey), Ranses José Vázquez (Cuba), Mark Weltz (The United States of America)

Contributing Authors: Mansour Almazroui (Saudi Arabia), Hamda Aloui (Tunisia), Hesham El-Askary (Egypt), Abdul Rasul Awan (Pakistan), Céline Bellard (France), Arden Burrell (Australia), Stefan van der Esch (The Netherlands), Robyn Hetem (South Africa), Kathleen Hermans (Germany), Margot Hurlbert (Canada), Jagdish Krishnaswamy (India), Zaneta Kubik (Poland), German Kust (The Russian Federation), Eike Lüdeling (Germany), Johan Meijer (The Netherlands), Ali Mohammed (Egypt), Katerina Michaelides (Cyprus/United Kingdom), Lindsay Stringer (United Kingdom), Stefan Martin Strohmeier (Austria), Grace Villamor (The Philippines)

Review Editors: Mariam Akhtar-Schuster (Germany), Fatima Driouech (Morocco), Mahesh Sankaran (India)

Chapter Scientists: Chuck Chuan Ng (Malaysia), Helen Berga Paulos (Ethiopia)

Date of Draft: 25/04/2019

| | | |
|----|---|----|
| 1 | Table of Contents | |
| 2 | | |
| 3 | Chapter 3 : Desertification | 1 |
| 4 | Executive Summary | 3 |
| 5 | 3.1. The Nature of Desertification | 7 |
| 6 | 3.1.1. Introduction | 7 |
| 7 | 3.1.2. Desertification in previous IPCC and related reports | 10 |
| 8 | 3.1.3. Dryland Populations: Vulnerability and Resilience | 11 |
| 9 | 3.1.4. Processes and Drivers of Desertification under Climate Change | 13 |
| 10 | 3.2. Observations of Desertification | 16 |
| 11 | 3.2.1. Status and Trends of Desertification | 16 |
| 12 | 3.2.2. Attribution of Desertification | 23 |
| 13 | 3.3. Desertification Feedbacks to Climate | 28 |
| 14 | 3.3.1. Sand and Dust Aerosols | 28 |
| 15 | 3.3.2. Changes in Surface Albedo | 30 |
| 16 | 3.3.3. Changes in Vegetation and Greenhouse Gas Fluxes | 30 |
| 17 | 3.4. Desertification Impacts on Natural and Socio-Economic Systems under Climate Change | 31 |
| 18 | 3.4.1. Impacts on Natural and Managed Ecosystems | 31 |
| 19 | 3.4.2. Impacts on Socio-economic Systems | 34 |
| 20 | 3.5. Future Projections | 40 |
| 21 | 3.5.1. Future Projections of Desertification | 40 |
| 22 | 3.5.2. Future Projections of Impacts | 42 |
| 23 | 3.6. Responses to Desertification under Climate Change | 44 |
| 24 | 3.6.1. SLM Technologies and Practices: on the Ground Actions | 45 |
| 25 | 3.6.2. Socio-economic Responses | 51 |
| 26 | 3.6.3. Policy Responses | 53 |
| 27 | Cross-Chapter Box 5: Policy Responses to Drought | 60 |
| 28 | 3.6.4. Limits to Adaptation, Maladaptation, and Barriers for Mitigation | 62 |
| 29 | 3.7. Hotspots and Case Studies | 63 |
| 30 | 3.7.1. Climate Change and Soil Erosion | 63 |
| 31 | 3.7.2. Green Walls and Green Dams | 67 |
| 32 | 3.7.3. Invasive Plant Species | 71 |
| 33 | 3.7.4. Oases in Hyper-arid Areas in the Arabian Peninsula and Northern Africa | 75 |
| 34 | 3.7.5. Integrated Watershed Management | 78 |
| 35 | 3.8. Knowledge Gaps and Key Uncertainties | 82 |
| 36 | Frequently Asked Questions | 83 |
| 37 | References | 84 |

1

2 **Executive Summary**

3 **Desertification is land degradation in arid, semi-arid, and dry sub-humid areas, collectively**
4 **known as drylands, resulting from many factors, including human activities and climatic**
5 **variations. The range and intensity of desertification have increased in some dryland areas over**
6 **the past several decades (*high confidence*).** Drylands currently cover about 46.2% ($\pm 0.8\%$) of the
7 global land area and are home to 3 billion people. The multiplicity and complexity of the processes of
8 desertification make its quantification difficult. Desertification hotspots, as identified by a decline in
9 vegetation productivity between 1980s and 2000s, extended to about 9.2% of drylands ($\pm 0.5\%$),
10 affecting about 500 (± 120) million people in 2015. The highest numbers of people affected are in
11 South and East Asia, North Africa and Middle East (*low confidence*). Desertification has already
12 reduced agricultural productivity and incomes (*high confidence*) and contributed to the loss of
13 biodiversity in some dryland regions (*medium confidence*). In many dryland areas, spread of invasive
14 plants has led to losses in ecosystem services (*high confidence*), while over-extraction is leading to
15 groundwater depletion (*high confidence*). Unsustainable land management, particularly when coupled
16 with droughts, has contributed to higher dust storm activity, reducing human wellbeing in drylands
17 and beyond (*high confidence*). Dust storms were associated with global cardiopulmonary mortality of
18 about 402,000 people in a single year. Higher intensity of sand storms and sand dune movements are
19 causing disruption and damage to transportation and solar and wind energy harvesting infrastructures
20 (*high confidence*). {3.1.1, 3.1.4, 3.2.1, 3.3.1, 3.4.1, 3.4.2, 3.4.2, 3.7.3, 3.7.4}

21 **Attribution of desertification to climate variability and change and human activities varies in**
22 **space and time (*high confidence*).** Climate variability and anthropogenic climate change, particularly
23 through increases in both land surface air temperature and evapotranspiration, and decreases in
24 precipitation, are *likely* to have played a role, in interaction with human activities, in causing
25 desertification in some dryland areas. The major human drivers of desertification interacting with
26 climate change are expansion of croplands, unsustainable land management practices and increased
27 pressure on land from population and income growth. Poverty is limiting both capacities to adapt to
28 climate change and availability of financial resources to invest in sustainable land management (SLM)
29 (*high confidence*). {3.1.4, 3.2.2, 3.4.2}

30 **Climate change will exacerbate several desertification processes (*medium confidence*).** Although
31 CO₂-fertilisation effect is enhancing vegetation productivity in drylands (*high confidence*), decreases
32 in water availability have a larger effect than CO₂-fertilisation in many dryland areas. There is *high*
33 *confidence* that aridity will increase in some places, but no evidence for a projected global trend in
34 dryland aridity (*medium confidence*). The area at risk of salinisation is projected to increase in the
35 future (*limited evidence, high agreement*). Future climate change is projected to increase the potential
36 for water driven soil erosion in many dryland areas (*medium confidence*), leading to soil organic
37 carbon decline in some dryland areas. {3.1.1, 3.2.2, 3.5.1, 3.5.2, 3.7.1, 3.7.3}

38 **Risks from desertification are projected to increase due climate change (*high confidence*).** Under
39 shared socioeconomic pathway SSP2 (“Middle of the Road”) at 1.5°C, 2°C and 3°C of global
40 warming, the number of dryland population exposed (vulnerable) to various impacts related to water,
41 energy and land sectors (e.g. water stress, drought intensity, habitat degradation) are projected to
42 reach 951 (178) million, 1,152 (220) million and 1,285 (277) million, respectively. While at global
43 warming of 2°C, under SSP1 (sustainability), the exposed (vulnerable) dryland population is 974 (35)
44 million, and under SSP3 (Fragmented World) it is 1,267 (522) million. Around half of the vulnerable
45 population is in South Asia, followed by Central Asia, West Africa and East Asia. {2.2, 3.1.1, 3.2.2,
46 3.5.1, 3.5.2, 7.2.2}

1 **Desertification and climate change, both individually and in combination, will reduce the**
2 **provision of dryland ecosystem services and lower ecosystem health, including losses in**
3 **biodiversity (*high confidence*).** Desertification and changing climate are projected to cause
4 reductions in crop and livestock productivity (*high confidence*), modify the composition of plant
5 species and reduce biological diversity across drylands (*medium confidence*). Rising CO₂ levels will
6 favour more rapid expansion of some invasive plant species in some regions. A reduction in the
7 quality and quantity of resources available to herbivores can have knock-on consequences for
8 predators, which can potentially lead to disruptive ecological cascades (*limited evidence, low*
9 *agreement*). Projected increases in temperature and the severity of drought events across some
10 dryland areas can increase chances of wildfire occurrence (*medium confidence*). {3.1.4, 3.4.1, 3.5.2,
11 3.7.3}

12 **Increasing human pressures on land combined with climate change will reduce the resilience of**
13 **dryland populations and constrain their adaptive capacities (*medium confidence*).** The
14 combination of pressures coming from climate variability, anthropogenic climate change and
15 desertification will contribute to poverty, food insecurity, and increased disease burden (*high*
16 *confidence*), as well as potentially to conflicts (*low confidence*). Although strong impacts of climate
17 change on migration in dryland areas are disputed (*medium evidence, low agreement*), in some places,
18 desertification under changing climate can provide an added incentive to migrate (*medium*
19 *confidence*). Women will be impacted more than men by environmental degradation, particularly in
20 those areas with higher dependence on agricultural livelihoods (*medium evidence, high agreement*).
21 {3.4.2, 3.6.2}

22 **Desertification exacerbates climate change through several mechanisms such as changes in**
23 **vegetation cover, sand and dust aerosols and greenhouse gas fluxes (*high confidence*).** The
24 **extent of areas in which dryness controls CO₂ exchange (rather than temperature) has increased**
25 **by 6% between 1948-2012, and is projected to increase by at least another 8% by 2050 if the**
26 **expansion continues at the same rate. In these areas, net carbon uptake is about 27% lower than**
27 **in other areas (*low confidence*).** Desertification also tends to increase albedo, decreasing energy
28 available at the surface and associated surface temperatures, producing a negative feedback on climate
29 change (*high confidence*). Through its effect on vegetation and soils, desertification changes the
30 absorption and release of associated greenhouse gases (GHGs). Vegetation loss and drying of surface
31 cover due to desertification increases the frequency of dust storms (*high confidence*). Arid ecosystems
32 could be an important global carbon sink depending on soil water availability (*medium evidence, high*
33 *agreement*). {3.3.3, 3.4.1, 3.5.2}

34 **Site-specific technological solutions, based both on new scientific innovations and indigenous**
35 **and local knowledge (ILK), are available to avoid, reduce and reverse desertification,**
36 **simultaneously contributing to climate change mitigation and adaptation (*high confidence*).**
37 SLM practices in drylands increase agricultural productivity and contribute to climate change
38 adaptation and mitigation (*high confidence*). Integrated crop, soil and water management measures
39 can be employed to reduce soil degradation and increase the resilience of agricultural production
40 systems to the impacts of climate change (*high confidence*). These measures include crop
41 diversification and adoption of drought-tolerant crops, reduced tillage, adoption of improved irrigation
42 techniques (e.g. drip irrigation) and moisture conservation methods (e.g. rainwater harvesting using
43 indigenous and local practices), and maintaining vegetation and mulch cover. Conservation
44 agriculture increases the capacity of agricultural households to adapt to climate change (*high*
45 *confidence*) and can lead to increases in soil organic carbon over time, with quantitative estimates of
46 the rates of carbon sequestration in drylands following changes in agricultural practices ranging
47 between 0.04-0.4 t ha⁻¹(*medium confidence*). Rangeland management systems based on sustainable

1 grazing and re-vegetation increase rangeland productivity and the flow of ecosystem services (*high*
2 *confidence*). The combined use of salt-tolerant crops, improved irrigation practices, chemical
3 remediation measures and appropriate mulch and compost is effective in reducing the impact of
4 secondary salinisation (*medium confidence*). Application of sand dune stabilisation techniques
5 contributes to reducing sand and dust storms (*high confidence*). Agroforestry practices and
6 shelterbelts help reduce soil erosion and sequester carbon. Afforestation programmes aimed at
7 creating windbreaks in the form of “green walls” and “green dams” can help stabilise and reduce dust
8 storms, avert wind erosion, and serve as carbon sinks, particularly when done with locally adapted
9 tree species (*high confidence*). {3.4.2, 3.6.1, 3.7.2}

10 **Investments into SLM, land restoration and rehabilitation in dryland areas have positive**
11 **economic returns (*high confidence*)**. Each USD invested into land restoration can have social returns
12 of about 3–6 USD over a 30-year period. Most SLM practices can become financially profitable
13 within three to 10 years (*medium evidence, high agreement*). Despite their benefits in addressing
14 desertification, mitigating and adapting to climate change, and increasing food and economic security,
15 many SLM practices are not widely adopted due to insecure land tenure, lack of access to credit and
16 agricultural advisory services, and insufficient incentives for private land users (*robust evidence, high*
17 *agreement*). {3.6.3}

18 **Indigenous and local knowledge (ILK) often contribute to enhancing resilience against climate**
19 **change and combating desertification (*medium confidence*)**. Dryland populations have developed
20 traditional agroecological practices which are well adapted to resource-sparse dryland environments.
21 However, there is *robust evidence* documenting losses of traditional agroecological knowledge.
22 Traditional agroecological practices are also increasingly unable to cope with growing demand for
23 food. Combined use of ILK and new SLM technologies can contribute to raising the resilience to the
24 challenges of climate change and desertification (*high confidence*). {3.1.3, 3.6.1, 3.6.2}

25 **Policy frameworks promoting the adoption of SLM solutions contribute to addressing**
26 **desertification as well as mitigating and adapting to climate change, with co-benefits for poverty**
27 **reduction and food security among dryland populations (*high confidence*)**. **Implementation of**
28 **Land Degradation Neutrality policies allows to avoid, reduce and reverse desertification, thus,**
29 **contributing to climate change adaptation and mitigation (*high confidence*)**. Strengthening land
30 tenure security is a major factor contributing to the adoption of soil conservation measures in
31 croplands (*high confidence*). On-farm and off-farm livelihood diversification strategies increase the
32 resilience of rural households against desertification and extreme weather events, such as droughts
33 (*high confidence*). Strengthening collective action is important for addressing causes and impacts of
34 desertification, and for adapting to climate change (*medium confidence*). A greater emphasis on
35 understanding gender-specific differences over land use and land management practices can help
36 make land restoration projects more successful (*medium confidence*). Improved access to markets
37 raises agricultural profitability and motivates investment into climate change adaptation and SLM
38 (*medium confidence*). Payments for ecosystem services give additional incentives to land users to
39 adopt SLM practices (*medium confidence*). Expanding access to rural advisory services increases the
40 knowledge on SLM and facilitates their wider adoption (*medium confidence*). Transition to modern
41 renewable energy sources can contribute to reducing desertification and mitigating climate change
42 through decreasing the use of fuelwood and crop residues for energy (*medium confidence*). Policy
43 responses to droughts based on pro-active drought preparedness and drought risk mitigation are more
44 efficient in limiting drought-caused damages than reactive drought relief efforts (*high confidence*).
45 {3.4.2, 3.6.2, 3.6.3, Cross-Chapter Box 5 in this chapter}

46 **The knowledge on limits to adaptation to combined effects of climate change and desertification**
47 **is insufficient. However, the potential for residual risks and maladaptive outcomes is high (*high***
48 ***confidence*)**. Empirical evidence on the limits to adaptation in dryland areas is limited, potential limits

1 to adaptation include losses of land productivity due to irreversible forms of desertification. Residual
2 risks can emerge from the inability of SLM measures to fully compensate for yield losses due to
3 climate change impacts, as well as foregone reductions in ecosystem services due to soil fertility loss
4 even when applying SLM measures could revert land to initial productivity after some time. Some
5 activities favouring agricultural intensification in dryland areas can become maladaptive due to their
6 negative impacts on the environment (*medium confidence*) {3.6.4}.

7 **Improving capacities, providing higher access to climate services, including local level early**
8 **warning systems, and expanding the use of remote sensing technologies are high return**
9 **investments for enabling effective adaptation and mitigation responses that help address**
10 **desertification (*high confidence*)**. Reliable and timely climate services, relevant to desertification,
11 can aid the development of appropriate adaptation and mitigation options reducing the impact of
12 desertification on human and natural systems (*high confidence*), with quantitative estimates pointing
13 that every USD invested in strengthening hydro-meteorological and early warning services in
14 developing countries can yield between 4 to 35 USD (*low confidence*). Knowledge and flow of
15 knowledge on desertification is currently fragmented. Improved knowledge and data exchange and
16 sharing will increase the effectiveness of efforts to achieve Land Degradation Neutrality (*high*
17 *confidence*). Expanded use of remotely sensed information for data collection helps in measuring
18 progress towards achieving Land Degradation Neutrality (*low evidence, high agreement*). {3.2.1,
19 3.6.2, 3.6.3, Cross-Chapter Box 5 in this chapter}

20

3.1. The Nature of Desertification

3.1.1. Introduction

In this report, desertification is defined as land degradation in arid, semi-arid, and dry sub-humid areas resulting from many factors, including climatic variations and human activities (United Nations Convention to Combat Desertification (UNCCD 1994). Land degradation is a negative trend in land condition, caused by direct or indirect human-induced processes including anthropogenic climate change, expressed as long-term reduction or loss of at least one of the following: biological productivity, ecological integrity or value to humans (4.1.3). Arid, semi-arid, and dry sub-humid areas, together with hyper-arid areas, constitute drylands (UNEP, 1992), home to about 3 billion people (van der Esch et al., 2017). The difference between desertification and land degradation is not process-based but geographic. Although land degradation can occur anywhere across the world, when it occurs in drylands, it is considered desertification (FAQ 1.3). Desertification is not limited to irreversible forms of land degradation, nor is it equated to desert expansion, but represents all forms and levels of land degradation occurring in drylands.

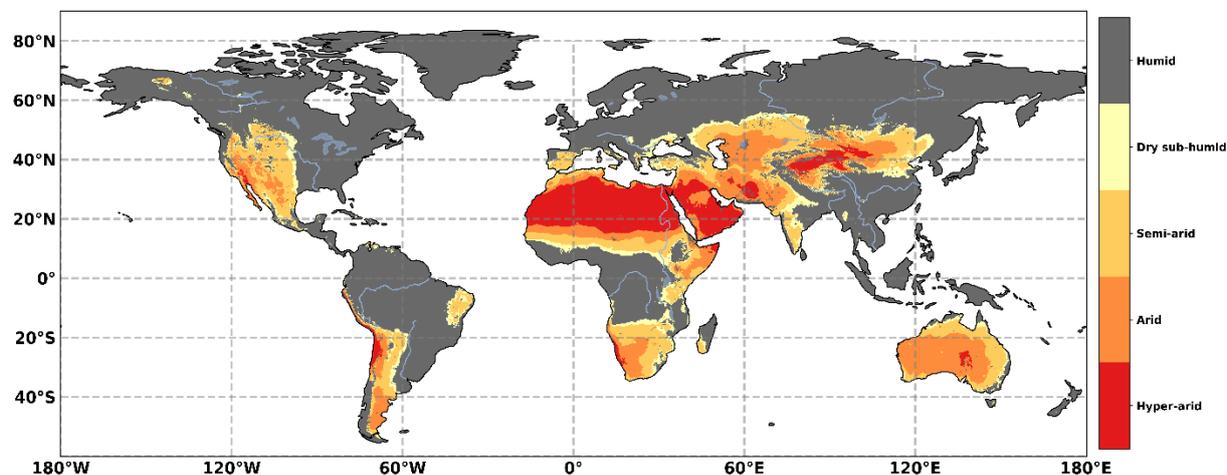
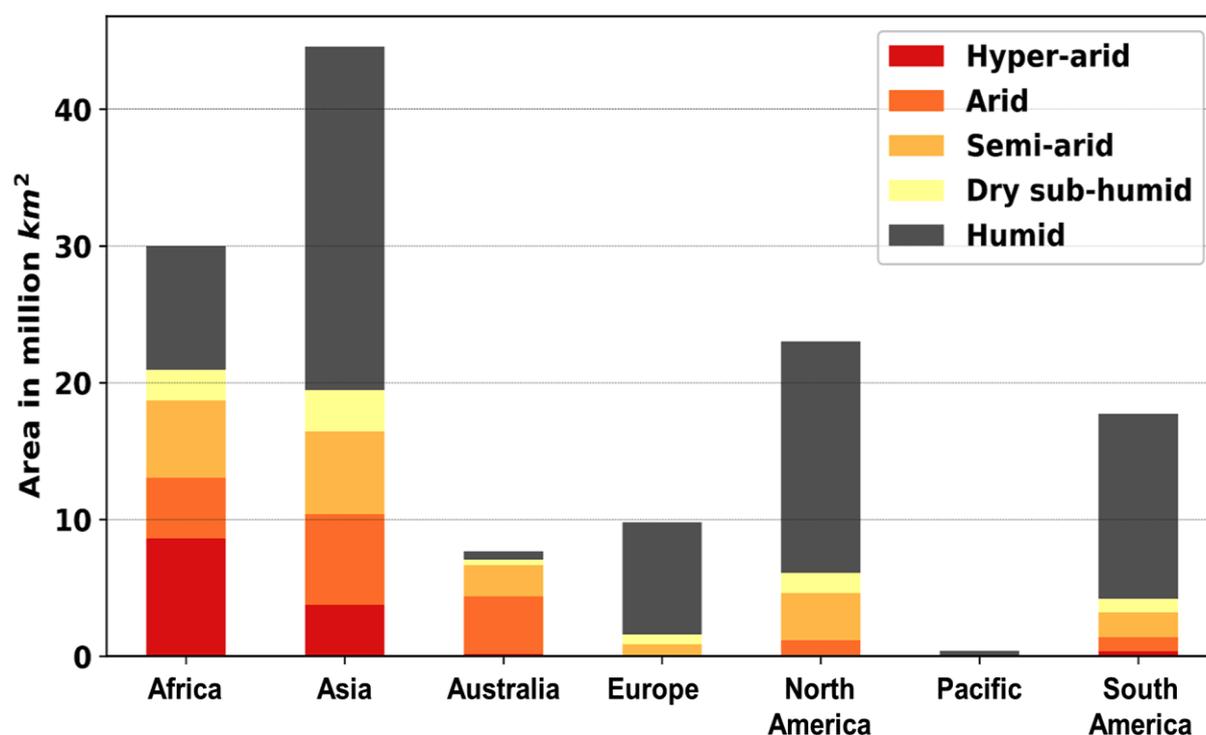


Figure 3.1 Geographical distribution of drylands, delimited based on the Aridity Index (AI). The classification of AI is: Humid $AI > 0.65$, Dry sub-humid $0.50 < AI \leq 0.65$, Semi-arid $0.20 < AI \leq 0.50$, Arid $0.05 < AI \leq 0.20$, Hyper-arid $AI < 0.05$. Data: TerraClimate precipitation and potential evapotranspiration (1980-2015) (Abatzoglou et al., 2018).

The geographic classification of drylands is often based on the aridity index (AI) - the ratio of average annual precipitation amount (P) to potential evapotranspiration amount (PET, see glossary) (Figure 3.1). Recent estimates, based on AI, suggest that drylands cover about 46.2% ($\pm 0.8\%$) of the global land area (Koutroulis, 2019; Právělie, 2016) (*low confidence*). Hyper-arid areas, where the aridity index is below 0.05, are included in drylands, but are excluded from the definition of desertification (UNCCD, 1994). Deserts are valuable ecosystems (UNEP, 2006; Safriel, 2009) geographically located in drylands and vulnerable to climate change. However, they are not considered prone to desertification. Aridity is a long-term climatic feature characterised by low average precipitation or available water (Gbeckor-Kove, 1989; Türkeş, 1999). Thus, aridity is different from drought which is a temporary climatic event (Maliva and Missimer, 2012). Moreover, droughts are not restricted to drylands, but occur both in drylands and humid areas (Wilhite et al., 2014). Following the Synthesis Report (SYR) of the IPCC Fifth Assessment Report (AR5), drought is defined here as “a period of abnormally dry weather long enough to cause a serious hydrological imbalance” (Mach et al., 2014; Cross-Chapter Box 5: Case study on policy responses to drought, in this chapter).

1 AI is not an accurate proxy for delineating drylands in an increasing CO₂ environment (3.2.1). The
 2 suggestion that most of the world has become more arid, since the AI has decreased, is not supported
 3 by changes observed in precipitation, evaporation or drought (Sheffield et al., 2012; Greve et al.,
 4 2014). While climate change is expected to decrease the AI due to increases in potential evaporation,
 5 the assumptions that underpin the potential evaporation calculation are not consistent with a changing
 6 CO₂ environment and the effect this has on transpiration rates (3.2.1; Roderick et al., 2015; Milly and
 7 Dunne, 2016; Greve et al., 2017). Given that future climate is characterised by significant increases in
 8 CO₂, the usefulness of currently applied AI thresholds to estimate dryland areas is limited under
 9 climate change. If instead of the AI, other variables such as precipitation, soil moisture, and primary
 10 productivity are used to identify dryland areas, there is no clear indication that the extent of drylands
 11 will change overall under climate change (Roderick et al., 2015; Greve et al., 2017; Lemordant et al.,
 12 2018). Thus, some dryland borders will expand, while some others will contract (*high confidence*).

13 Approximately 70% of dryland areas are located in Africa and Asia (Figure 3.2). The biggest land
 14 use/cover in terms of area in drylands, if deserts are excluded, are grasslands, followed by forests and
 15 croplands (Figure 3.3). The category of “other lands” in Figure 3.3 includes bare soil, ice, rock, and
 16 all other land areas that are not included within the other five categories (FAO, 2016). Thus, hyper-
 17 arid areas contain mostly deserts, with some small exceptions, for example, where grasslands and
 18 croplands are cultivated under oasis conditions with irrigation (3.7.4). Moreover, FAO (2016) defines
 19 grasslands as permanent pastures and meadows used continuously for more than five years. In
 20 drylands, transhumance, i.e. seasonal migratory grazing, often leads to non-permanent pasture
 21 systems, thus, some of the areas under “other land” category are also used as non-permanent pastures
 22 (Ramankutty et al., 2008; Fetzel et al., 2017; Erb et al., 2016).

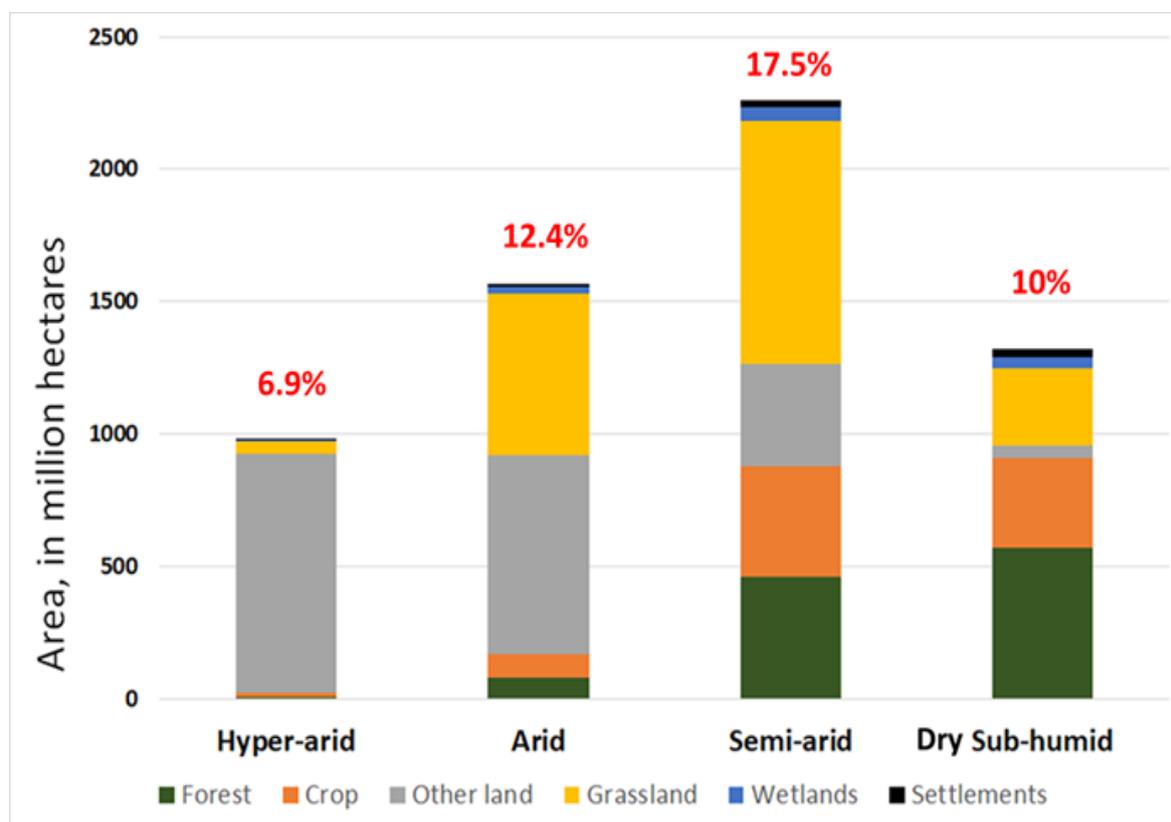


23

24 **Figure 3.2 Dryland categories across geographical areas (continents and Pacific region). Data:**
 25 **TerraClimate precipitation and potential evapotranspiration (1980-2015) (Abatzoglou et al., 2018).**

26 In the earlier global assessments of desertification (since the 1970s), which were based on qualitative
 27 expert evaluations, the extent of desertification was found to range between 4% and 70% of the area
 28 of drylands (Safriel, 2007). More recent estimates, based on remotely sensed data, show that about
 29 24–29% of the global land area experienced reductions in biomass productivity between 1980s and

1 2000s (Bai et al., 2008; Le et al., 2016), corresponding to about 9.2% of drylands ($\pm 0.5\%$)
 2 experiencing declines in biomass productivity during this period (*low confidence*), mainly due to
 3 anthropogenic causes. Both of these studies consider rainfall dynamics, thus, accounting for the effect
 4 of droughts. While less than 10% of drylands is undergoing desertification, it is occurring in areas that
 5 contain around 20% of dryland population (Klein Goldewijk et al., 2017). In these areas the population
 6 has increased from ~172 million in 1950 to over 630 million today (Figure 1.1).



7
 8 **Figure 3.3 Land use and land cover in drylands and share of each dryland category in global land area.**
 9 **Source: FAO (2016).**

10 Available assessments of the global extent and severity of desertification are relatively crude
 11 approximations with considerable uncertainties, for example, due to confounding effects of invasive
 12 bush encroachment in some dryland regions. Different indicator sets and approaches have been
 13 developed for monitoring and assessment of desertification from national to global scales (Imeson,
 14 2012; Sommer et al., 2011; Zucca et al., 2012; Bestelmeyer et al., 2013). Many indicators of
 15 desertification only include a single factor or characteristic of desertification, such as the patch size
 16 distribution of vegetation (Maestre and Escudero, 2009; Kéfi et al., 2010), Normalized Difference
 17 Vegetation Index (NDVI) (Piao et al., 2005), drought-tolerant plant species (An et al., 2007), grass
 18 cover (Bestelmeyer et al., 2013), land productivity dynamics (Baskan et al., 2017), ecosystem net
 19 primary productivity (Zhou et al., 2015) or environmentally sensitive land area index (Symeonakis et
 20 al., 2016). In addition, some synthetic indicators of desertification have also been used to assess
 21 desertification extent and desertification processes, such as climate, land use, soil, and socioeconomic
 22 parameters (Dharumarajan et al., 2018), or changes in climate, land use, vegetation cover, soil
 23 properties and population as the desertification vulnerability index (Salvati et al., 2009). Current data
 24 availability and methodological challenges do not allow for accurately and comprehensively mapping
 25 desertification at a global scale (Cherlet et al., 2018). However, the emerging partial evidence points
 26 to a lower global extent of desertification than previously estimated (*medium confidence*) (3.2).

1 This assessment examines the socio-ecological links between drivers (3.1) and feedbacks (3.3) that
2 influence desertification-climate change interactions, and then examines associated observed and
3 projected impacts (3.4, 3.5) and responses (3.6). Moreover, this assessment highlights that dryland
4 populations are highly vulnerable to desertification and climate change (3.2, 3.4). At the same time,
5 dryland populations also have significant past experience and sources of resilience embodied in
6 indigenous and local knowledge and practices in order to successfully adapt to climatic changes and
7 address desertification (3.6). Numerous site-specific technological response options are also available
8 for SLM in drylands that can help increase the resilience of agricultural livelihood systems to climate
9 change (3.6). However, continuing environmental degradation combined with climate change are
10 straining the resilience of dryland populations. Enabling policy responses for SLM and livelihoods
11 diversification can help maintain and strengthen the resilience and adaptive capacities in dryland areas
12 (3.6). The assessment finds that policies promoting SLM in drylands will contribute to climate change
13 adaptation and mitigation, with co-benefits for broader sustainable development (*high confidence*)
14 (3.4).

16 3.1.2. Desertification in previous IPCC and related reports

17 The IPCC Fifth Assessment report (AR5) and Special Report on Global Warming of 1.5°C include a
18 limited discussion of desertification. In AR5 Working Group I desertification is mentioned as a
19 forcing agent for the production of atmospheric dust (Myhre et al., 2013). The same report had *low*
20 *confidence* in the available projections on the changes in dust loadings due to climate change
21 (Boucher et al., 2013). In AR5 Working Group II, desertification is identified as a process that can
22 lead to reductions in crop yields and the resilience of agricultural and pastoral livelihoods (Field et al.,
23 2014; Klein et al., 2015). AR5 Working Group II notes that climate change will amplify water
24 scarcity with negative impacts on agricultural systems, particularly in semi-arid environments of
25 Africa (*high confidence*), while droughts could exacerbate desertification in south-western parts of
26 Central Asia (Field et al., 2014). AR5 Working Group III identifies desertification as one of a number
27 of often overlapping issues that must be dealt with when considering governance of mitigation and
28 adaptation (Fleurbaey et al., 2014). The IPCC Special Report on Global Warming of 1.5°C noted that
29 limiting global warming to 1.5°C instead of 2°C is strongly beneficial for land ecosystems and their
30 services (*high confidence*) such as soil conservation, contributing to avoidance of desertification
31 (Hoegh-Guldberg et al., 2018).

32 The recent Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services
33 (IPBES) Assessment report on land degradation and restoration (IPBES, 2018a) is also of particular
34 relevance. While acknowledging a wide variety of past estimates of the area undergoing degradation,
35 IPBES (2018a) pointed at their lack of agreement about where degradation is taking place. IPBES
36 (2018a) also recognised the challenges associated with differentiating the impacts of climate
37 variability and change on land degradation from the impacts of human activities at a regional or
38 global scale.

39 The third edition of the World Atlas of Desertification (Cherlet et al., 2018) indicated that it is not
40 possible to deterministically map the global extent of land degradation, and its subset - desertification,
41 pointing out that the complexity of interactions between social, economic, and environmental systems
42 make land degradation not amenable to mapping at a global scale. Instead, Cherlet et al. (2018)
43 presented global maps highlighting the convergence of various pressures on land resources.

3.1.3. Dryland Populations: Vulnerability and Resilience

Drylands are home to approximately 38.2% ($\pm 0.6\%$) of the global population (Koutroulis, 2019; van der Esch et al., 2017), that is about 3 billion people. The highest number of people live in the drylands of South Asia (Figure 3.4), followed by Sub-Saharan Africa and Latin America (van der Esch et al., 2017). In terms of the number of people affected by desertification, Reynolds et al. (2007) indicated that desertification was directly affecting 250 million people. More recent estimates show that 500 (± 120) million people lived in 2015 in those dryland areas which experienced significant loss in biomass productivity between 1980s and 2000s (Bai et al., 2008; Le et al., 2016). The highest numbers of affected people were in South and East Asia, North Africa and Middle East (*low confidence*). The population in drylands is projected to increase about twice as rapidly as non-drylands, reaching 4 billion people by 2050 (van der Esch et al., 2017). This is due to higher population growth rates in drylands. About 90% of the population in drylands live in developing countries (UN-EMG, 2011).

Dryland populations are highly vulnerable to desertification and climate change (Howe et al., 2013; Huang et al., 2016, 2017; Liu et al., 2016b; Thornton et al., 2014; Lawrence et al., 2018) because their livelihoods are predominantly dependent on agriculture; one of the sectors most susceptible to climate change (Rosenzweig et al., 2014; Schlenker and Lobell, 2010). Climate change is projected to have substantial impacts on all types of agricultural livelihood systems in drylands (CGIAR-RPDS, 2014) (3.4.1, 3.4.2).

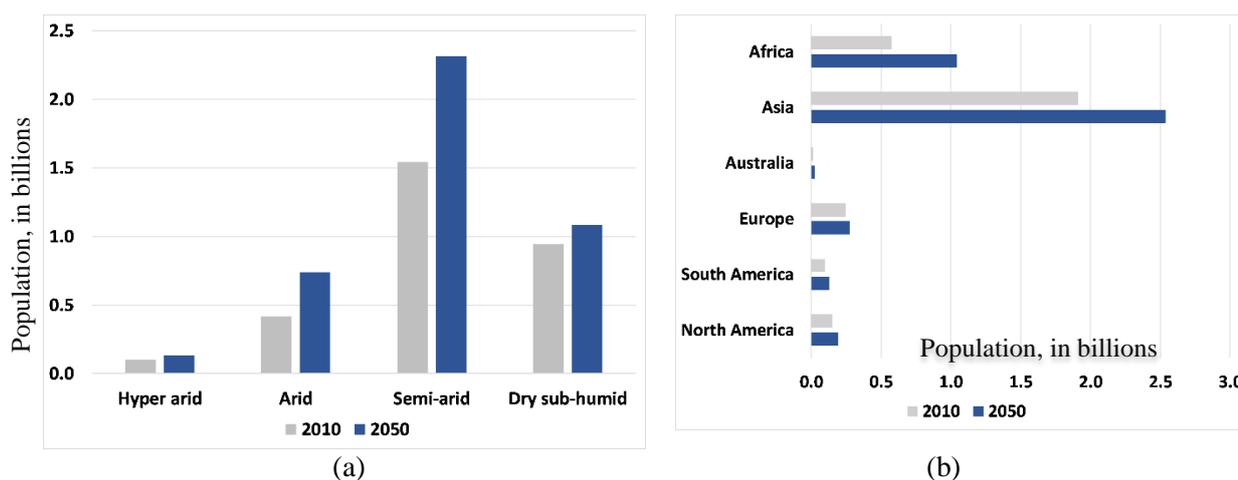


Figure 3.4 Current (a) and projected population (under SSP2) (b) in drylands, in billions.

Source: van der Esch et al. (2017)

One key vulnerable group in drylands are pastoral and agropastoral households¹. There are no precise figures about the number of people practicing pastoralism globally. Most estimates range between 100 to 200 million (Rass, 2006; Secretariat of the Convention on Biological Diversity, 2010), of whom 30–63 million are nomadic pastoralists (Dong, 2016; Carr-Hill, 2013)². Pastoral production systems represent an adaptation to high seasonal climate variability and low biomass productivity in dryland ecosystems (Varghese and Singh, 2016; Krätli and Schareika, 2010), which require large areas for

¹FOOTNOTE: Pastoralists derive more than 50% of their income from livestock and livestock products, whereas agro-pastoralists generate more than 50% of their income from crop production and at least 25% from livestock production (Swift, 1988).

²FOOTNOTE: The estimates of the number of pastoralists, and especially of nomadic pastoralists, are very uncertain, because often nomadic pastoralists are not fully captured in national surveys and censuses (Carr-Hill, 2013).

1 livestock grazing through migratory pastoralism (Snorek et al., 2014). Grazing lands across dryland
2 environments are being degraded, and/or being converted to crop production, limiting the
3 opportunities for migratory livestock systems, and leading to conflicts with sedentary crop producers
4 (Abbass, 2014; Dimelu et al., 2016). These processes, coupled with ethnic differences, perceived
5 security threats, and misunderstanding of pastoral rationality, have led to increasing marginalisation
6 of pastoral communities and disruption of their economic and cultural structures (Elhadary, 2014;
7 Morton, 2010). As a result, pastoral communities are not well prepared to deal with increasing
8 weather/climate variability and weather/climate extremes due to changing climate (Dong, 2016;
9 López-i-Gelats et al., 2016), and remain amongst the most food insecure groups in the world (FAO,
10 2018).

11 There is an increasing concentration of poverty in the dryland areas of Sub-Saharan Africa and South
12 Asia (von Braun and Gatzweiler, 2014; Barbier and Hochard, 2016), where 41% and 12% of the total
13 populations live in extreme poverty, respectively (World Bank, 2018). For comparison, the average
14 share of global population living in extreme poverty is about 10% (World Bank, 2018).
15 Multidimensional poverty, prevalent in many dryland areas, is a key source of vulnerability (Safriel et
16 al., 2005; Thornton et al., 2014; Fraser et al., 2011; Thomas, 2008). Multidimensional poverty
17 incorporates both income-based poverty, and also other dimensions such as poor healthcare services,
18 lack of education, lack of access to water, sanitation and energy, disempowerment, and threat from
19 violence (Bourguignon and Chakravarty, 2003; Alkire and Santos, 2010, 2014). Contributing
20 elements to this multidimensional poverty in drylands are rapid population growth, fragile
21 institutional environment, lack of infrastructure, geographic isolation and low market access, insecure
22 land tenure systems, and low agricultural productivity (Sietz et al., 2011; Reynolds et al., 2011;
23 Safriel and Adeel, 2008; Stafford Smith, 2016). Even in high-income countries, those dryland areas
24 that depend on agricultural livelihoods represent relatively poorer locations nationally, with fewer
25 livelihood opportunities, for example in Italy (Salvati, 2014). Moreover, in many drylands areas,
26 female-headed households, women and subsistence farmers (both male and female) are more
27 vulnerable to the impacts of desertification and climate change (Nyantakyi-Frimpong and Bezner-
28 Kerr, 2015; Sultana, 2014; Rahman, 2013). Some local cultural traditions and patriarchal relationships
29 were found to contribute to higher vulnerability of women and female-headed households through
30 restrictions on their access to productive resources (Nyantakyi-Frimpong and Bezner-Kerr, 2015;
31 Sultana, 2014; Rahman, 2013) (3.4.2, 3.6.3; Cross-Chapter Box 11: Gender, Chapter 7).

32 Despite these environmental, socio-economic and institutional constraints, dryland populations have
33 historically demonstrated remarkable resilience, ingenuity and innovations, distilled into indigenous
34 and local knowledge to cope with high climatic variability and sustain livelihoods (Safriel and Adeel,
35 2008; Davis, 2016; Davies, 2017; 3.6.1, 3.6.2; Cross-Chapter Box 13: Indigenous and Local
36 Knowledge, Chapter 7). For example, across the Arabian Peninsula and North Africa, informal
37 community bylaws were successfully used for regulating grazing, collection and cutting of herbs and
38 wood, that limited rangeland degradation (Gari, 2006; Hussein, 2011). Pastoralists in Mongolia
39 developed indigenous classifications of pasture resources which facilitated ecologically optimal
40 grazing practices (Fernandez-Gimenez, 2000) (3.6.2). Currently, however, indigenous and local
41 knowledge and practices are increasingly lost or can no longer cope with growing demands for land-
42 based resources (Dominguez, 2014; Fernández-Giménez and Fillat Estaque, 2012; Hussein, 2011;
43 Kodirekkala, 2017; Moreno-Calles et al., 2012; 3.4.2). Unsustainable land management is increasing
44 the risks from droughts, floods and dust storms (3.4.2, 3.5). Policy actions promoting the adoption of
45 SLM practices in dryland areas, based on both indigenous and local knowledge and modern science,
46 and expanding alternative livelihood opportunities outside agriculture can contribute to climate
47 change adaptation and mitigation, addressing desertification, with co-benefits for poverty reduction
48 and food security (*high confidence*) (Cowie et al., 2018; Liniger et al., 2017; Safriel and Adeel, 2008;
49 Stafford-Smith et al., 2017).

3.1.4. Processes and Drivers of Desertification under Climate Change

3.1.4.1 Processes of Desertification and Their Climatic Drivers

Processes of desertification are mechanisms by which drylands are degraded. Desertification consists of both biological and non-biological processes. These processes are classified under broad categories of degradation of physical, chemical and biological properties of terrestrial ecosystems. The number of desertification processes is large and they are extensively covered elsewhere (IPBES, 2018a; Lal, 2016; Racine, 2008; UNCCD, 2017). Section 4.2.1 and Tables 4.1-4.2 in Chapter 4 highlight those which are particularly relevant for this assessment in terms of their links to climate change and land degradation, including desertification.

Drivers of desertification are factors which trigger desertification processes. Initial studies of desertification during the early-to-mid 20th century attributed it entirely to human activities. In one of the influential publications of that time, Lavauden (1927) stated that: "Desertification is purely artificial. It is only the act of the man..." However, such a uni-causal view on desertification was shown to be invalid (Geist et al., 2004; Reynolds et al., 2007) (3.1.4.2, 3.1.4.3). Tables 4.1-4.2 in Chapter 4 summarise drivers, linking them to the specific processes of desertification and land degradation under changing climate.

Erosion refers to removal of soil by the physical forces of water, wind, or often caused by farming activities such as tillage (Ginoux et al., 2012). The global estimates of soil erosion differ significantly, depending on scale, study period and method used (García-Ruiz et al., 2015), ranging from approximately 20 Gt yr⁻¹ to more than 200 Gt yr⁻¹ (Boix-Fayos et al., 2006; FAO, 2015). There is a significant potential for climate change to increase soil erosion by water particularly in those regions where precipitation volumes and intensity are projected to increase (Panthou et al., 2014; Nearing et al., 2015). On the other hand, while it is a dominant form of erosion in areas such as West Asia and the Arabian Peninsula (Prakash et al., 2015; Klingmüller et al., 2016), there is *limited evidence* concerning climate change impacts on wind erosion (Tables 4.1-4.2 in Chapter 4; 3.5).

Saline and sodic soils (see glossary) occur naturally in arid, semiarid and dry sub-humid regions of the world. Climate change or hydrological change can cause soil salinisation by increasing the mineralised ground water level. However, secondary salinisation occurs when the concentration of dissolved salts in water and soil is increased by anthropogenic processes, mainly through poorly managed irrigation schemes. The threat of soil and groundwater salinisation induced by sea level rise and sea water intrusion are amplified by climate change (4.9.7).

Global warming is expected to accelerate soil organic carbon (SOC) turnover, since the decomposition of the soil organic matter by microbial activity begins with low soil water availability, but this moisture is insufficient for plant productivity (Austin et al., 2004; 3.4.1.1), as well as losses by soil erosion (Lal, 2009); therefore, in some dryland areas leading to SOC decline (3.3.3; 3.5.2) and the transfer of carbon (C) from soil to the atmosphere (Lal, 2009).

Sea surface temperature (SST) anomalies can drive rainfall changes, with implications for desertification processes. North Atlantic SST anomalies are positively correlated with Sahel rainfall anomalies (Knight et al., 2006; Gonzalez-Martin et al., 2014; Sheen et al., 2017). While the eastern tropical Pacific SST anomalies have a negative correlation with Sahel rainfall (Pomposi et al., 2016), a cooler north Atlantic is related to a drier Sahel, with this relationship enhanced if there is a simultaneous relative warming of the south Atlantic (Hoerling et al., 2006). Huber and Fensholt (2011) explored the relationship between SST anomalies and satellite observed Sahel vegetation dynamics finding similar relationships but with substantial west-east variations in both the significant SST regions and the vegetation response. Concerning the paleoclimatic evidence on aridification after the early Holocene "Green Sahara" period (11,000 to 5000 years ago), Tierney et al. (2017) indicate

1 that a cooling of the north Atlantic played a role (Collins et al., 2017; Otto-Bliesner et al., 2014;
2 Niedermeyer et al., 2009) similar to that found in modern observations. Besides these SST
3 relationships, aerosols have also been suggested as a potential driver of the Sahel droughts (Rotstayn
4 and Lohmann, 2002; Booth et al., 2012; Ackerley et al., 2011). For Eastern Africa, both recent
5 droughts and decadal declines have been linked to human-induced warming in the western Pacific
6 (Funk et al., 2018).

7 Invasive plants contributed to desertification and loss of ecosystem services in many dryland areas in
8 the last century (*high confidence*) (3.7.3). Extensive woody plant encroachment altered runoff and soil
9 erosion across much of the drylands, because the bare soil between shrubs is very susceptible to water
10 erosion, mainly in high-intensity rainfall events (Manjoro et al., 2012; Pierson et al., 2013; Eldridge et
11 al., 2015). Rising CO₂ levels due to global warming favour more rapid expansion of some invasive
12 plant species in some regions. An example is the Great Basin region in western North America where
13 over 20% of ecosystems have been significantly altered by invasive plants, especially exotic annual
14 grasses and invasive conifers resulting in loss of biodiversity. This land cover conversion has resulted
15 in reductions in forage availability, wildlife habitat, and biodiversity (Pierson et al., 2011, 2013;
16 Miller et al., 2013).

17 The wildfire is a driver of desertification, because it reduces vegetation cover, increases runoff and
18 soil erosion, reduces soil fertility and affects the soil microbial community (Vega et al., 2005; Nyman
19 et al., 2010; Holden et al., 2013; Pourreza et al., 2014; Weber et al., 2014; Liu and Wimberly, 2016).
20 Predicted increases in temperature and the severity of drought events across some dryland areas (2.2)
21 can increase chances of wildfire occurrence (*medium confidence*) (Jolly et al., 2015; Williams et al.,
22 2010; Clarke and Evans, 2018; Cross-Chapter Box 3: Fire and Climate Change, Chapter 2). In
23 semiarid and dry sub-humid areas, fire can have a profound influence on observed vegetation and
24 particularly the relative abundance of grasses to woody plants (Bond et al., 2003; Bond and Keeley,
25 2005; Balch et al., 2013).

26 While large uncertainty exists concerning trends in droughts globally (AR5, 2.2), examining the
27 drought data by Ziese et al. (2014) for drylands only reveals a large inter-annual variability combined
28 with a trend toward increasing dryland area affected by droughts since 1950s (Figure 1.1).

29 **3.1.4.2. Anthropogenic Drivers of Desertification under Climate Change**

30 The literature on the human drivers of desertification is substantial (D'Odorico et al., 2013; Sietz et
31 al., 2011; Yan and Cai, 2015; Sterk et al., 2016; Varghese and Singh, 2016; to list a few) and there
32 have been several comprehensive reviews and assessments of these drivers very recently (Cherlet et
33 al., 2018; IPBES, 2018a; UNCCD, 2017). IPBES (2018a) identified cropland expansion,
34 unsustainable land management practices including overgrazing by livestock, urban expansion,
35 infrastructure development, and extractive industries as the main drivers of land degradation. IPBES
36 (2018a) also found that the ultimate driver of land degradation is high and growing consumption of
37 land-based resources, e.g. through deforestation and cropland expansion, escalated by population
38 growth. What is particularly relevant in the context of the present assessment is to evaluate if, how
39 and which human drivers of desertification will be modified by climate change effects.

40 Growing food demand is driving conversion of forests, rangelands, and woodlands into cropland
41 (Bestelmeyer et al., 2015; D'Odorico et al., 2013). Climate change is projected to reduce crop yields
42 across dryland areas (3.4.1; 5.2.2), potentially reducing local production of food and feed. Without
43 research breakthroughs mitigating these productivity losses through higher agricultural productivity,
44 and reducing food waste and loss, meeting increasing food demands of growing populations will
45 require expansion of cropped areas to more marginal areas (with most prime areas in drylands already
46 being under cultivation) (Lambin, 2012; Lambin et al., 2013; Eitelberg et al., 2015; Gutiérrez-Elorza,
47 2006; Kapović Solomun et al., 2018). Borrelli et al. (2017) showed that the primary driver of soil
48 erosion in 2012 was cropland expansion. Although local food demands could also be met by

1 importing from other areas, this would mean increasing the pressure on land in those areas (Lambin
2 and Meyfroidt, 2011). The net effects of such global agricultural production shifts on land condition
3 in drylands are not known.

4 Climate change will exacerbate poverty among some categories of dryland populations (3.4.2; 3.5.2).
5 Depending on the context, this impact comes through declines in agricultural productivity, changes in
6 agricultural prices and extreme weather events (Hertel and Lobell, 2014; Hallegatte and Rozenberg,
7 2017). There is *high confidence* that poverty limits both capacities to adapt to climate change and
8 availability of financial resources to invest into SLM (3.5.2; 3.6.2; 3.6.3; Gerber et al., 2014; Way,
9 2016; Vu et al., 2014).

10 Labour mobility is another key human driver which will interact with climate change. Although
11 strong impacts of climate change on migration in dryland areas are disputed, in some places, it is
12 *likely* to provide an added incentive to migrate (3.4.2.7). Out-migration will have several
13 contradictory effects on desertification. On one hand, it reduces an immediate pressure on land if it
14 leads to less dependence on land for livelihoods (Chen et al., 2014; Liu et al., 2016a). Moreover,
15 migrant remittances could be used to fund the adoption of SLM practices. Labour mobility from
16 agriculture to non-agricultural sectors could allow land consolidation, gradually leading to
17 mechanisation and agricultural intensification (Wang et al., 2014, 2018). On the other hand, this can
18 increase the costs of labour-intensive SLM practices due to lower availability of rural agricultural
19 labour and/or higher rural wages. Out-migration increases the pressure on land if higher wages that
20 rural migrants earn in urban centres will lead to their higher food consumption. Moreover, migrant
21 remittances could also be used to fund land use expansion to marginal areas (Taylor et al., 2016; Gray
22 and Bilsborrow, 2014). The net effect of these opposite mechanisms varies from place to place (Qin
23 and Liao, 2016). There is very little literature evaluating these joint effects of climate change,
24 desertification and labour mobility (7.3.2).

25 There are also many other institutional, policy and socio-economic drivers of desertification, such as
26 land tenure insecurity, lack of property rights, lack of access to markets, and to rural advisory
27 services, lack of technical knowledge and skills, agricultural price distortions, agricultural support and
28 subsidies contributing to desertification, and lack of economic incentives for SLM (D’Odorico et al.,
29 2013; Geist et al., 2004; Moussa et al., 2016; Mythili and Goedecke, 2016; Sow et al., 2016; Tun et
30 al., 2015; García-Ruiz, 2010). There is no evidence that these factors will be materially affected by
31 climate change, however, serving as drivers of unsustainable land management practices, they do play
32 a very important role in modulating responses for climate change adaptation and mitigation (3.6.3).

33 ***3.1.4.3 Interaction of Drivers: Desertification Syndrome versus Drylands Development*** 34 ***Paradigm***

35 Two broad narratives have historically emerged to describe responses of dryland populations to
36 environmental degradation. The first is “desertification syndrome” which describes the vicious cycle
37 of resource degradation and poverty, whereby dryland populations apply unsustainable agricultural
38 practices leading to desertification, and exacerbating their poverty, which then subsequently further
39 limits their capacities to invest in SLM (MEA, 2005; Safriel and Adeel, 2008). The alternative
40 paradigm is one of “drylands development”, which refers to social and technical ingenuity of dryland
41 populations as a driver of dryland sustainability (MEA, 2005; Reynolds et al., 2007; Safriel and
42 Adeel, 2008). The major difference between these two frameworks is that the “drylands development
43 paradigm” recognises that human activities are not the sole and/or most important drivers of
44 desertification, but there are interactions of human and climatic drivers within coupled social-
45 ecological systems (Reynolds et al., 2007). This led Behnke and Mortimore (2016), and earlier Swift
46 (1996), to conclude that the concept of desertification as irreversible degradation distorts policy and
47 governance in the dryland areas. Mortimore (2016) suggested that instead of externally imposed
48 technical solutions, what is needed is for populations in dryland areas to adapt to this variable

1 environment which they cannot control. All in all, there is *high confidence* that anthropogenic and
2 climatic drivers interact in complex ways in causing desertification. As discussed in Section 3.2.2, the
3 relative influence of human or climatic drivers on desertification varies from place to place (*high*
4 *confidence*) (Bestelmeyer et al., 2018; D’Odorico et al., 2013; Geist and Lambin, 2004; Kok et al.,
5 2016; Polley et al., 2013; Ravi et al., 2010; Scholes, 2009; Sietz et al., 2017; Sietz et al., 2011).

7 **3.2. Observations of Desertification**

8 **3.2.1. Status and Trends of Desertification**

9 Current estimates of the extent and severity of desertification vary greatly due to missing and/or
10 unreliable information (Gibbs and Salmon, 2015). The multiplicity and complexity of the processes of
11 desertification make its quantification difficult (Prince, 2016; Cherlet et al., 2018). The most common
12 definition for the drylands is based on defined thresholds of the AI (Figure 3.1) (UNEP, 1992). While
13 past studies have used the AI to examine changes in desertification or extent of the drylands (Feng
14 and Fu, 2013; Zarch et al., 2015; Ji et al., 2015; Spinoni et al., 2015; Huang et al., 2016; Ramarao et
15 al., 2018), this approach has several key limitations: (i) the AI does not measure desertification, (ii)
16 the impact of changes in climate on the land surface and systems is more complex than assumed by
17 AI, and (iii) the relationship between climate change and changes in vegetation is complex due to the
18 influence of CO₂. Expansion of the drylands does not imply desertification by itself, if there is no
19 long-term loss of at least one of the following: biological productivity, ecological integrity, and value
20 to humans.

21 The use of the AI to define changing aridity levels and dryland extent in an environment with
22 changing atmospheric CO₂ has been strongly challenged (Roderick et al., 2015; Milly and Dunne,
23 2016; Greve et al., 2017; Liu et al., 2017). The suggestion that most of the world has become more
24 arid, since the AI has decreased, is not supported by changes observed in precipitation, evaporation or
25 drought (Sheffield et al., 2012; Greve et al., 2014) (*medium confidence*). A key issue is the
26 assumption in the calculation of potential evapotranspiration that stomatal conductance remains
27 constant which is invalid if atmospheric CO₂ changes. Given that atmospheric CO₂ has been
28 increasing over the last century or more, and is projected to continue increasing, this means that AI
29 with constant thresholds (or any other measure that relies on potential evapotranspiration) is not an
30 appropriate way to estimate aridity or dryland extent (Donohue et al., 2013; Roderick et al., 2015;
31 Greve et al., 2017). This issue helps explain the apparent contradiction between the drylands
32 becoming more arid according to the AI and also becoming greener according to satellite observations
33 (Fensholt et al., 2012; Andela et al., 2013; Figure 3.5). Other climate type classifications based on
34 various combinations of temperature and precipitation (Köppen-Trewartha, Köppen-Geiger) have also
35 been used to examine historical changes in climate zones finding a tendency toward drier climate
36 types (Feng et al., 2014; Spinoni et al., 2015).

37 The need to establish a baseline when assessing change in the land area degraded has been extensively
38 discussed in Prince et al. (2018). Desertification is a process not a state of the system, hence an
39 “absolute” baseline is not required; however, every study uses a baseline defined by the start of their
40 period of interest.

41 Depending on the definitions applied and methodologies used in evaluation, the status and extent of
42 desertification globally and regionally still show substantial variations (D’Odorico et al., 2013) (*high*
43 *confidence*). There is *high confidence* that the range and intensity of desertification has increased in
44 some dryland areas over the past several decades (3.2.1.1, 3.2.1.2). The three methodological
45 approaches applied for assessing the extent of desertification: expert judgement, satellite observation
46 of net primary productivity, and use of biophysical models, together provide a relatively holistic

1 assessment but none on its own captures the whole picture (Gibbs and Salmon, 2015; Vogt et al.,
2 2011; Prince, 2016; 4.2.4).

3 3.2.1.1. Global Scale

4 Complex human-environment interactions coupled with biophysical, social, economic and political
5 factors unique to any given location render desertification difficult to map at a global scale (Cherlet et
6 al., 2018). Early attempts to assess desertification focused on expert knowledge in order to obtain
7 global coverage in a cost-effective manner. **Expert judgement** continues to play an important role
8 because degradation remains a subjective feature whose indicators are different from place to place
9 (Sonneveld and Dent, 2007). GLASOD (Global Assessment of Human-Induced Soil Degradation)
10 estimated nearly 2000 million hectares (M ha) (15.3% of the total land area) had been degraded by
11 early 1990s since mid-20th century. GLASOD was criticised for perceived subjectiveness and
12 exaggeration (Helldén and Tottrup, 2008; Sonneveld and Dent, 2007). Dregne and Chou (1992) found
13 3000 M ha in drylands (i.e. about 50% of drylands) were undergoing degradation. Significant
14 improvements have been made through the efforts of WOCAT (World Overview of Conservation
15 Approaches and Technologies), LADA (Land Degradation Assessment in Drylands) and DESIRE
16 (Desertification Mitigation and Remediation of Land) who jointly developed a mapping tool for
17 participatory expert assessment, with which land experts can estimate current area coverage, type and
18 trends of land degradation (Reed et al., 2011).

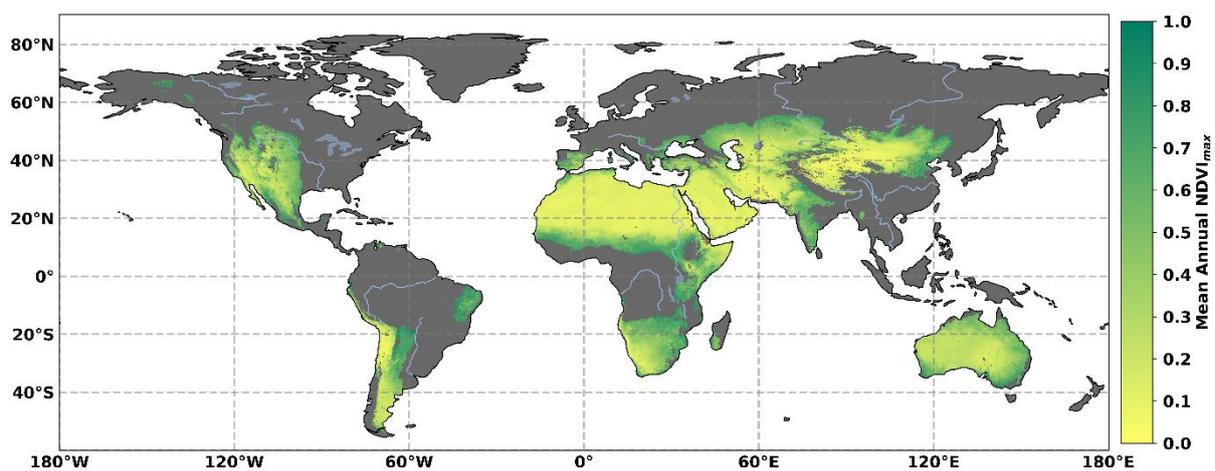
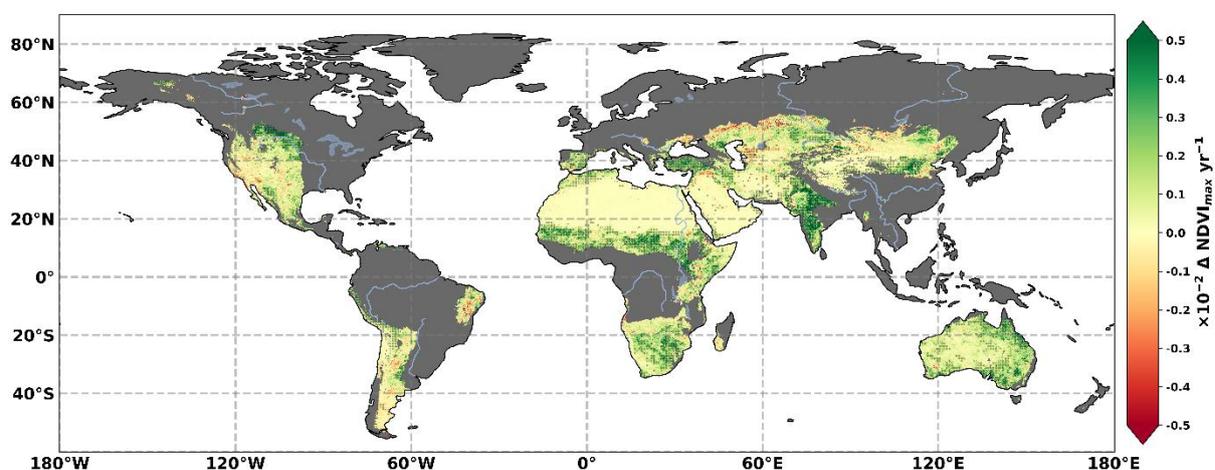


Figure 3.5 Mean Annual Maximum NDVI 1982-2015 (Global Inventory Modelling and Mapping Studies NDVI3g v1). Non-dryland regions (Aridity Index > 0.65) are masked in grey.



1 **Figure 3.6 Trend in the Annual Maximum NDVI 1982-2015 (Global Inventory Modelling and Mapping**
2 **Studies NDVI3g v1) calculated using the Theil-Sen estimator which is a median based estimator, and is**
3 **robust to outliers. Non-dryland regions (Aridity Index > 0.65) are masked in grey.**

4 A number of studies have used **satellite-based remote sensing** to investigate long-term changes in
5 the vegetation and thus identify parts of the drylands undergoing desertification. Satellite data
6 provides information at the resolution of the sensor which can be relatively coarse (up to 25 km) and
7 interpretations of the data at sub-pixel levels are challenging. The most widely used remotely sensed
8 vegetation index is the NDVI providing a measure of canopy greenness, which is related to the
9 quantity of standing biomass (Bai et al., 2008; de Jong et al., 2011; Fensholt et al., 2012; Andela et al.,
10 2013; Fensholt et al., 2015; Le et al., 2016; Figure 3.5). A main challenge associated with NDVI is
11 that although biomass and productivity are closely related in some systems, they can differ widely
12 when looking across land uses and ecosystem types, giving a false positive in some instances
13 (Pattison et al., 2015; Aynekulu et al., 2017). For example, bush encroachment in rangelands and
14 intensive monocropping with high fertiliser application gives an indication of increased productivity
15 in satellite data though these could be considered as land degradation. According to this measure there
16 are regions undergoing desertification, however, the drylands are greening on average (Figure 3.6).

17 A simple linear trend in NDVI is an unsuitable measure for dryland degradation for several reasons
18 (Wessels et al., 2012; de Jong et al., 2013; Higginbottom and Symeonakis, 2014; Le et al., 2016).
19 NDVI is strongly coupled to precipitation in drylands where precipitation has high inter-annual
20 variability. This means that NDVI trend can be dominated by any precipitation trend and is sensitive
21 to wet or dry periods, particularly if they fall near the beginning or end of the time series. Degradation
22 may only occur during part of the time series, while NDVI is stable or even improving during the rest
23 of the time series. This reduces the strength and representativeness of a linear trend. Other factors
24 such as CO₂ fertilisation also influence the NDVI trend. Various techniques have been proposed to
25 address these issues, including the residual trends (RESTREND) method to account for rainfall
26 variability (Evans and Geerken, 2004), time-series break point identification methods to find major
27 shifts in the vegetation trends (de Jong et al., 2013; Verbesselt et al., 2010a) and methods to explicitly
28 account for the effect of CO₂ fertilisation (Le et al., 2016).

29 Using the RESTREND method, Andela et al. (2013) found that human activity contributed to a
30 mixture of improving and degrading regions in drylands. In some locations these regions differed
31 substantially from those identified using the NDVI trend alone, including an increase in the area being
32 desertified in southern Africa and northern Australia, and a decrease in southeast and west Australia
33 and Mongolia. De Jong et al. (2013) examined the NDVI time series for major shifts in vegetation
34 activity and found that 74% of drylands experienced such a shift between 1981 and 2011. This
35 suggests that monotonic linear trends are unsuitable for accurately capturing the changes that have
36 occurred in the majority of the drylands. Le et al. (2016) explicitly accounted for CO₂ fertilisation
37 effect and found that the extent of degraded areas in the world is 3% larger when compared to the
38 linear NDVI trend.

39 Besides NDVI, there are many vegetation indices derived from satellite data in the optical and
40 infrared wavelengths. Each of these datasets has been derived to overcome some limitation in existing
41 indices. Studies have compared vegetation indices globally (Zhang et al., 2017) and specifically over
42 drylands (Wu, 2014). In general, the data from these vegetation indices are available only since
43 around 2000, while NDVI data is available since 1982. With less than 20 years of data, the trend
44 analysis remains problematic with vegetation indices other than NDVI. However, given the various
45 advantages in terms of resolution and other characteristics, these newer vegetation indices will
46 become more useful in the future as more data accumulates.

47 Vegetation Optical Depth (VOD) has been available since the 1980s. VOD is based on microwave
48 measurements and is related to total above ground biomass water content. Unlike NDVI which is only

1 sensitive to green canopy cover, VOD is also sensitive to water in woody parts of the vegetation and
2 hence provides a view of vegetation changes that can be complementary to NDVI. Liu et al. (2013)
3 used VOD trends to investigate biomass changes and found that VOD was closely related to
4 precipitation changes in drylands. To complement their work with NDVI, Andela et al. (2013) also
5 applied the RESTREND method to VOD. By interpreting NDVI and VOD trends together they were
6 able to differentiate changes to the herbaceous and woody components of the biomass. They reported
7 that many dryland regions are experiencing an increase in the woody fraction often associated with
8 shrub encroachment and suggest that this was aided by CO₂ fertilisation.

9 A major shortcoming of these studies based on vegetation datasets derived from satellite sensors is
10 that they do not account for changes in vegetation composition, thus leading to inaccuracies in the
11 estimation of the extent of degraded areas in drylands. For example, drylands of Eastern Africa
12 currently face growing encroachment of invasive plant species, such as *Prosopis juliflora* (Ayanu et
13 al., 2015), which constitutes land degradation since it leads to losses in economic productivity of
14 affected areas but appears as a greening in the satellite data. Another case study in central Senegal
15 found degradation manifested through a reduction in species richness despite satellite observed
16 greening (Herrmann and Tappan, 2013). A number of efforts to identify changes in vegetation
17 composition from satellites have been made (Brandt et al., 2016a,b; Evans and Geerken, 2006;
18 Geerken, 2009; Geerken et al., 2005; Verbesselt et al., 2010a,b). These depend on well-identified
19 reference NDVI time series for particular vegetation groupings, can only differentiate vegetation types
20 that have distinct spectral phenology signatures and require extensive ground observations for
21 validation. A recent alternative approach to differentiating woody from herbaceous vegetation
22 involves the combined use of optical/infrared based vegetation indices, indicating greenness, with
23 microwave based Vegetation Optical Depth (VOD) which is sensitive to both woody and leafy
24 vegetation components (Andela et al., 2013; Tian et al., 2017).

25 **Biophysical models** use global data sets that describe climate patterns and soil groups, combined with
26 observations of land use, to define classes of potential productivity and map general land degradation
27 (Gibbs and Salmon, 2015). All biophysical models have their own set of assumptions and limitations
28 that contribute to their overall uncertainty, including: model structure; spatial scale; data requirements
29 (with associated errors); spatial heterogeneities of socioeconomic conditions; and agricultural
30 technologies used. Models have been used to estimate the vegetation productivity potential of land
31 (Cai et al., 2011) and to understand the causes of observed vegetation changes. Zhu et al. (2016) used
32 an ensemble of ecosystem models to investigate causes of vegetation changes from 1982-2009, using
33 a factorial simulation approach. They found CO₂ fertilisation to be the dominant effect globally
34 though climate and land cover change were the dominant effects in various dryland locations. Borrelli
35 et al. (2017) modelled that about 6.1% of the global land area experienced very high soil erosion rates
36 (exceeding 10 Mg ha⁻¹ yr⁻¹) in 2012, particularly in South America, Africa, and Asia.

37 Overall, improved estimation and mapping of areas undergoing desertification are needed. This
38 requires a combination of rapidly expanding sources of remotely sensed data, ground observations
39 and new modelling approaches. This is a critical gap, especially in the context of measuring progress
40 towards achieving the land degradation-neutrality target by 2030 in the framework of SDGs.

41 3.2.1.2. Regional Scale

42 While global scale studies provide information for any region, there are numerous studies that focus
43 on sub-continental scales, providing more in-depth analysis and understanding. Regional and local
44 studies are important to detect location-specific trends in desertification and heterogeneous influences
45 of climate change on desertification. However, these regional and local studies use a wide variety of
46 methodologies, making direct comparisons difficult. For details of the methodologies applied by each
47 study refer to the individual papers.

1 3.2.1.2.1 Africa

2 It is estimated that 46 out of the 54 countries in Africa are vulnerable to desertification, with some
3 already affected (Právālie, 2016). Moderate or higher severity degradation over recent decades have
4 been identified in many river basins including the Nile (42% of area), Niger (50%), Senegal (51%),
5 Volta (67%), Limpopo (66%) and Lake Chad (26%) (Thiombiano and Tourino-Soto, 2007).

6 The Horn of Africa is getting drier (Damberg and AghaKouchak, 2014; Marshall et al., 2012)
7 exacerbating the desertification already occurring (Oroda, 2001). The observed decline in vegetation
8 cover is diminishing ecosystem services (Pricope et al., 2013). Based on NDVI residuals, Kenya
9 experienced persistent negative (positive) trends over 21.6% (8.9%) of the country, for the period
10 1992–2015 (Gichenje and Godinho, 2018). Fragmentation of habitats, reduction in the range of
11 livestock grazing, higher stocking rates are considered to be the main drivers for vegetation structure
12 loss in the rangelands of Kenya (Kihuu, 2016; Otuoma et al., 2009)

13 Despite desertification in the Sahel being a major concern since the 1970s, wetting and greening
14 conditions have been observed in this region over the last three decades (Anyamba and Tucker, 2005;
15 Huber et al., 2011; Brandt et al., 2015; Rishmawi et al., 2016; Tian et al., 2016; Leroux et al., 2017;
16 Herrmann et al., 2005; Damberg and AghaKouchak, 2014). Cropland areas in the Sahel region of
17 West Africa have doubled since 1975, with settlement area increasing by about 150% (Traore et al.,
18 2014). Thomas and Nigam (2018) found that the Sahara expanded by 10% over the 20th century based
19 on annual rainfall. In Burkina Faso, Dimobe et al. (2015) estimated that from 1984 to 2013, bare soils
20 and agricultural lands increased by 18.8% and 89.7%, respectively, while woodland, gallery forest,
21 tree savannas, shrub savannas and water bodies decreased by 18.8%, 19.4%, 4.8%, 45.2% and 31.2%,
22 respectively. In Fakara region in Niger, 5% annual reduction in herbaceous yield between 1994 and
23 2006 was largely explained by changes in land use, grazing pressure and soil fertility (Hiernaux et al.,
24 2009). Aladejana et al. (2018) found that between 1986 and 2015, 18.6% of the forest cover around
25 the Owena River basin was lost. For the period 1982–2003, Le et al. (2012) found that 8% of the
26 Volta River basin's landmass had been degraded with this increasing to 65% after accounting for the
27 effects of CO₂ (+NO_x) fertilisation.

28 Greening has also been observed in parts of Southern Africa but it is relatively weak compared to
29 other regions of the continent (Helldén and Tottrup, 2008; Fensholt et al., 2012). However, greening
30 can be accompanied by desertification when factors such as decreasing species richness, changes in
31 species composition and shrub encroachment are observed (Smith et al., 2013; Herrmann and Tappan,
32 2013; Kaptué et al., 2015; Herrmann and Sop, 2016; Saha et al., 2015) (3.1.4, 3.7.3). In the Okavango
33 river Basin in Southern Africa, conversion of land towards higher utilisation intensities, unsustainable
34 agricultural practises and overexploitation of the savanna ecosystems have been observed in recent
35 decades (Weinzierl et al., 2016).

36 In arid Algerian High Plateaus, desertification due to both climatic and human causes led to the loss
37 of indigenous plant biodiversity between 1975 and 2006 (Hirche et al., 2011). Ayoub (1998)
38 identified 64 M ha in Sudan as degraded, with the Central North Kordofan state being most affected.
39 However, reforestation measures in the last decade sustained by improved rainfall conditions have led
40 to low-medium regrowth conditions in about 20% of the area (Dawelbait and Morari, 2012). In
41 Morocco, areas affected by desertification were dominantly on plains with high population and
42 livestock pressure (del Barrio et al., 2016; Kouba et al., 2018; Lahlaoui et al., 2017). The annual costs
43 of soil degradation were estimated at about 1% of Gross Domestic Product (GDP) in Algeria and
44 Egypt, and about 0.5% in Morocco and Tunisia (Réquier-Desjardins and Bied-Charretton, 2006).

45 3.2.1.2.2 Asia

46 Právālie (2016) found that desertification is currently affecting 38 of 48 countries in Asia. The
47 changes in drylands in Asia over the period 1982–2011 were mixed, with some areas experiencing
48 vegetation improvement while others showed reduced vegetation (Miao et al., 2015a). Major river

1 basins undergoing salinisation include: Indo-Gangetic Basin in India (Lal and Stewart, 2012), Indus
2 Basin in Pakistan (Aslam and Prathapar, 2006), Yellow River Basin in China (Chengrui and Dregne,
3 2001), Yinchuan Plain, in China (Zhou et al., 2013), Aral Sea Basin of Central Asia (Cai et al., 2003;
4 Pankova, 2016; Qadir et al., 2009).

5 Helldén and Tottrup (2008) highlighted a greening trend in East Asia between 1982 and 2003. Over
6 the past several decades, air temperature and the rainfall increased in the arid and hyper-arid region of
7 Northwest China (Chen et al., 2015; Wang et al., 2017). Within China, rainfall erosivity has shown a
8 positive trend in dryland areas between 1961 and 2012 (Yang and Lu, 2015). While water erosion
9 area in Xinjiang China, has decreased by 23.2%, erosion considered as severe or intense was still
10 increasing (Zhang et al., 2015). Xue et al. (2017) used remote sensing data covering 1975 to 2015 to
11 show that wind-driven desertified land in north Shanxi in China had expanded until 2000, before
12 contracting again. Li et al. (2012) used satellite data to identify desertification in Inner Mongolia
13 China and found a link between policy changes and the locations and extent of human-caused
14 desertification. Several oasis regions in China have seen increases in cropland area, while forests,
15 grasslands and available water resources have decreased (Fu et al. 2017; Muyibul et al., 2018; Xie et
16 al., 2014). Between 1990 and 2011 15.3% of Hognu Khaan nature reserve in central Mongolia was
17 subjected to desertification (Lamchin et al., 2016). Using satellite data Liu et al. (2013) found the area
18 of Mongolia undergoing non-climatic desertification was associated with increases in goat density and
19 wildfire occurrence.

20 In Central Asia, drying up of the Aral Sea is continuing having negative impacts on regional
21 microclimate and human health (Issanova and Abuduwaili, 2017; Lioubimtseva, 2015; Micklin, 2016;
22 Xi and Sokolik, 2015). Half of the region's irrigated lands, especially in the Amudarya and Syrdarya
23 river basins, were affected by secondary salinisation (Qadir et al., 2009). Le et al., (2016) showed that
24 about 57% of croplands in Kazakhstan and about 20% of croplands in Kyrgyzstan had lost in their
25 vegetation productivity between 1982 and 2006. Chen et al. (2019) indicated that about 58% of the
26 grasslands in the region lost in their vegetation productivity between 1999 and 2015. Anthropogenic
27 factors were the main driver of this loss in Turkmenistan and Uzbekistan, while the role of human
28 drivers was smaller than that of climate-related factors in Tajikistan and Kyrgyzstan (Chen et al.,
29 2019). The total costs of land degradation in Central Asia were estimated to equal about USD 6
30 billion annually (Mirzabaev et al., 2016).

31 Damberg and AghaKouchak (2014) found that parts of South Asia experienced drying over the last
32 three decades. More than 75% of the area of northern, western and southern Afghanistan is affected
33 by overgrazing and deforestation (UNEP-GEF, 2008). Desertification is a serious problem in Pakistan
34 with a wide range of human and natural causes (Irshad et al., 2007; Lal, 2018). Similarly,
35 desertification affects parts of India (Kundu et al., 2017; Dharumarajan et al., 2018; Christian et al.,
36 2018). Using satellite data to map various desertification processes, Ajai et al. (2009) identified 81.4
37 M ha were subject to various processes of desertification in India in 2005, while salinisation affected
38 6.73 M ha in the country (Singh, 2009).

39 Saudi Arabia is highly vulnerable to desertification (Ministry of Energy Industry and Mineral
40 Resources, 2016), with this vulnerability expected to increase in the north-western parts of the country
41 in the coming decades. Yahiya (2012) found that Jazan, south-western Saudi Arabia, lost about 46%
42 of its vegetation cover from 1987 to 2002. Droughts and frequent dust storms were shown to impose
43 adverse impacts over Saudi Arabia especially under global warming and future climate change
44 (Hasanean et al., 2015). In north-west Jordan, 18% of the area was prone to severe to very severe
45 desertification (Al-Bakri et al., 2016). Large parts of the Syrian drylands have been identified as
46 undergoing desertification (Evans and Geerken, 2004; Geerken and Ilaiwi, 2004). Moridnejad et al.
47 (2015) identified newly desertified regions in the Middle East based on dust sources, finding that

1 these regions accounted for 39% of all detected dust source points. Desertification has increased
2 substantially in Iran since the 1930s. Despite numerous efforts to rehabilitate degraded areas, it still
3 poses a major threat to agricultural livelihoods in the country (Amiraslani and Dragovich, 2011).

4 *3.2.1.2.3 Australia*

5 Damberg and AghaKouchak (2014) found that wetter conditions were experienced in northern
6 Australia over the last three decades with widespread greening observed between 1981 and 2006 over
7 much of Australia, except for eastern Australia where large areas were affected by droughts from
8 2002 to 2009 based on Advanced High Resolution Radiometer (AVHRR) satellite data (Donohue,
9 McVicar, and Roderick, 2009). For the period 1982–2013, Burrell et al. (2017) also found widespread
10 greening over Australia including eastern Australia over the post-drought period. This dramatic
11 change in the trend found for eastern Australia emphasises the dominant role played by precipitation
12 in the drylands. Degradation due to anthropogenic activities and other causes affects over 5% of
13 Australia, particularly near the central west coast. Jackson and Prince (2016) used a local NPP scaling
14 approach applied with MODIS derived vegetation data to quantify degradation in a dryland watershed
15 in Northern Australia from 2000 to 2013. They estimated that 20% of the watershed was degraded.
16 Salinisation has also been found to be degrading parts of the Murray-Darling Basin in Australia
17 (Rengasamy, 2006). Eldridge and Soliveres (2014) examined areas undergoing woody encroachment
18 in eastern Australia and found that rather than degrading the landscape, the shrubs often enhanced
19 ecosystem services.

20 *3.2.1.2.4 Europe*

21 Drylands cover 33.8% of northern Mediterranean countries; approximately 69% of Spain, 66% of
22 Cyprus, and between 16% and 62% in Greece, Portugal, Italy and France (Zdruli, 2011). The
23 European Environment Agency (EEA) indicated that 14 M ha, i.e. 8% of the territory of the European
24 Union (in Bulgaria, Cyprus, Greece, Italy, Romania, Spain and Portugal), had a “very high” and “high
25 sensitivity” to desertification (European Court of Auditors, 2018). This figure increases to 40 M ha
26 (23% of the EU territory) if “moderately” sensitive areas are included (Právělie et al., 2017; European
27 Court of Auditors, 2018). Desertification in the region is driven by irrigation developments and
28 encroachment of cultivation on rangelands (Safriel, 2009) caused by population growth, agricultural
29 policies and markets. According to a recent assessment report (ECA, 2018), Europe is increasingly
30 affected by desertification leading to significant consequences on land use, particularly in Portugal,
31 Spain, Italy, Greece, Malta, Cyprus, Bulgaria and Romania. Using the Universal Soil Loss Equation,
32 it was estimated that soil erosion can be as high as $300 \text{ t ha}^{-1}\text{yr}^{-1}$ (equivalent to a net loss of 18 mm yr^{-1})
33 in Spain (López-Bermúdez, 1990). For the badlands region in south-east Spain, however, it was
34 shown that biological soil crusts effectively prevent soil erosion (Lázaro et al., 2008). In
35 Mediterranean Europe, Guerra et al. (2016) found a reduction of erosion due to greater effectiveness
36 of soil erosion prevention between 2001 and 2013. Helldén and Tottrup (2008) observed a greening
37 trend in the Mediterranean between 1982–2003, while Fensholt et al. (2012) also show a dominance
38 of greening in Eastern Europe.

39 In Russia, at the beginning of the 2000s, about 7% of the total area (i.e. ~130 M ha) was threatened by
40 desertification (Gunin and Pankova, 2004; Kust et al., 2011). Turkey is considered highly vulnerable
41 to drought, land degradation and desertification (Türkeş, 1999; Türkeş, 2003). About 60% of Turkey’s
42 land area is characterised with hydro-climatological conditions favourable for desertification (Türkeş,
43 2013). ÇEMGM (2017) estimated that about half of Turkey’s land area (48.6%) is prone to moderate
44 to high desertification.

45 *3.2.1.2.5 North America*

46 Drylands cover approximately 60% of Mexico. According to Pontifes et al. (2018), 3.5% of the area
47 was converted from natural vegetation to agriculture and human settlements between 2002 to 2011.

1 The region is highly vulnerable to desertification due to frequent droughts and floods (Méndez and
2 Magaña, 2010; Stahle et al., 2009; Becerril-Pina Rocio et al., 2015).

3 For the period 2000-2011 the overall difference between potential and actual NPP in different land
4 capability classes in the south-western United States was 11.8% (Noojipady et al., 2015); reductions
5 in grassland-savanna and livestock grazing area and forests were the highest. Bush encroachment is
6 observed over a fairly wide area of western USA grasslands; including Jornada Basin within the
7 Chihuahuan Desert, and is spreading at a fast rate despite grazing restrictions intended to curb the
8 spread (Yanoff and Muldavin, 2008; Browning and Archer, 2011; Van Auken, 2009; Rachal et al.,
9 2012). In comparing sand dune migration patterns and rates between 1995 and 2014, Potter and
10 Weigand (2016) established that the area covered by stable dune surfaces, and sand removal zones,
11 decreased while sand accumulation zones increased from 15.4 to 25.5 km² for Palen dunes in
12 Southern California Desert, while movement of Kelso Dunes is less clear (Lam et al., 2011). Within
13 the United States, average soil erosion rates on all croplands decreased by about 38% between 1982-
14 2003 due to better soil management practices (Kertis, 2003).

15 3.2.1.2.6 *Central and South America*

16 Morales et al. (2011) indicated that desertification costs between 8 and 14% of gross agricultural
17 product in many Central and South American countries. Parts of the dry Chaco and Caldenal regions
18 in Argentina have undergone widespread degradation over the last century (Verón et al., 2017;
19 Fernández et al., 2009). Bisigato and Laphitz (2009) identified overgrazing as a cause of
20 desertification in the Patagonian Monte region of Argentina. Vieira et al. (2015) found that 94% of
21 northeast Brazilian drylands were susceptible to desertification. It is estimated that up to 50% of the
22 area was being degraded due to frequent prolonged droughts and clearing of forests for agriculture.
23 This land-use change threatens the extinction of around 28 native species (Leal et al., 2005). In
24 Central Chile, dryland forest and shrubland area was reduced by 1.7% and 0.7%, respectively,
25 between 1975-2008 (Schulz et al., 2010).

26

27 **3.2.2. Attribution of Desertification**

28 Desertification is a result of complex interactions within coupled social-ecological systems. Thus, the
29 relative contributions of climatic, anthropogenic and other drivers of desertification vary depending
30 on specific socioeconomic and ecological contexts. The high natural climate variability in dryland
31 regions is a major cause of vegetation changes but does not necessarily imply degradation. Drought is
32 not degradation as the land productivity may return entirely once the drought ends (Kassas, 1995).
33 However, if droughts increase in frequency, intensity and/or duration they may overwhelm the
34 vegetation's ability to recover (ecosystem resilience, Prince et al., 2018), causing degradation.
35 Assuming a stationary climate and no human influence, rainfall variability results in fluctuations in
36 vegetation dynamics which can be considered temporary as the ecosystem tends to recover with
37 rainfall, and desertification does not occur (Ellis, 1995; Vetter, 2005; von Wehrden et al., 2012).
38 Climate change on the other hand, exemplified by a non-stationary climate, can gradually cause a
39 persistent change in the ecosystem through aridification and CO₂ changes. Assuming no human
40 influence, this 'natural' climatic version of desertification may take place rapidly, especially when
41 thresholds are reached (Prince et al., 2018), or over longer periods of time as the ecosystems slowly
42 adjust to a new climatic norm through progressive changes in the plant community composition.
43 Accounting for this climatic variability is required before attributions to other causes of desertification
44 can be made.

45 For attributing vegetation changes to climate versus other causes, rain use efficiency (RUE - the
46 change in net primary productivity (NPP) per unit of precipitation) and its variations in time have
47 been used (Prince et al., 1998). Global applications of RUE trends to attribute degradation to climate
48 or other (largely human) causes has been performed by Bai et al. (2008) and Le et al. (2016) (3.2.1.1).

1 The RESTREND (residual trend) method analyses the correlation between annual maximum NDVI
2 (or other vegetation index as a proxy for NPP) and precipitation by testing accumulation and lag
3 periods for the precipitation (Evans and Geerken, 2004). The identified relationship with the highest
4 correlation represents the maximum amount of vegetation variability that can be explained by the
5 precipitation, and corresponding RUE values can be calculated. Using this relationship, the climate
6 component of the NDVI time series can be reconstructed, and the difference between this and the
7 original time series (the residual) is attributed to anthropogenic and other causes.

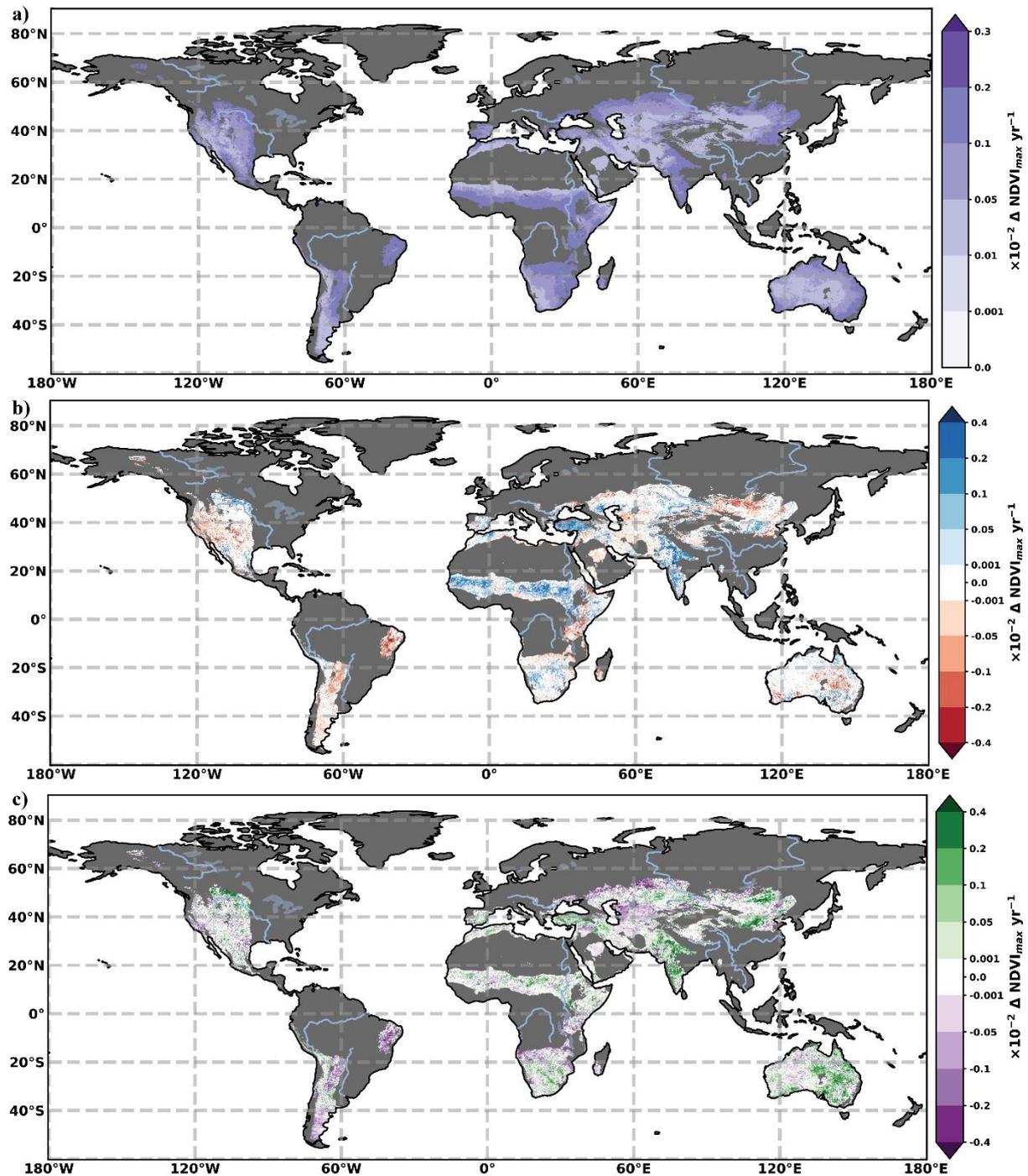
8 The RESTREND method, or minor variations of it, have been applied extensively. (Herrmann and
9 Hutchinson, 2005) concluded that climate was the dominant causative factor for widespread greening
10 in the Sahel region from 1982 to 2003, and anthropogenic and other factors were mostly producing
11 land improvements or no change. However, pockets of desertification were identified in Nigeria and
12 Sudan. Similar results were also found from 1982 to 2007 by Huber et al. (2011). Wessels et al.
13 (2007) applied RESTREND to South Africa and showed that RESTREND produced a more accurate
14 identification of degraded land than RUE alone. RESTREND identified a smaller area undergoing
15 desertification due to non-climate causes compared to the NDVI trends. Liu et al. (2013) extended the
16 climate component of RESTREND to include temperature and applied this to VOD observations of
17 the cold drylands of Mongolia. They found the area undergoing desertification due to non-climatic
18 causes is much smaller than the area with negative VOD trends. RESTREND has also been applied in
19 several other studies to the Sahel (Leroux et al., 2017), Somalia (Omuto et al., 2010), West Africa
20 (Ibrahim et al., 2015), China (Li et al., 2012; Yin et al., 2014), Central Asia (Jiang et al., 2017),
21 Australia (Burrell et al., 2017) and globally (Andela et al., 2013). In each of these studies the extent to
22 which desertification can be attributed to climate versus other causes varies across the landscape.

23 These studies represent the best regional, remote sensing based attribution studies to date, noting that
24 RESTREND and RUE have some limitations (Higginbottom and Symeonakis, 2014). Vegetation
25 growth (NPP) changes slowly compared to rainfall variations and may be sensitive to rainfall over
26 extended periods (years) depending on vegetation type. Detection of lags and the use of weighted
27 antecedent rainfall can partially address this problem though most studies do not do this. The method
28 addresses changes since the start of the time series, it cannot identify whether an area is already
29 degraded at the start time. It is assumed that climate, particularly rainfall, are principal factors in
30 vegetation change which may not be true in more humid regions.

31 Another assumption in RESTREND is that any trend is linear throughout the period examined. That
32 is, there are no discontinuities (break points) in the trend. Browning et al. (2017) have shown that
33 break points in NDVI time series reflect vegetation changes based on long-term field sites. To
34 overcome this limitation, Burrell et al. (2017) introduced the Time Series Segmentation-RESTREND
35 (TSS-RESTREND) which allows a breakpoint or turning point within the period examined (Figure
36 3.7). Using TSS-RESTREND over Australia they identified more than double the degrading area than
37 could be identified with a standard RESTREND analysis. The occurrence and drivers of abrupt
38 change (turning points) in ecosystem functioning were also examined by Horion et al. (2016) over the
39 semi-arid Northern Eurasian agricultural frontier. They combined trend shifts in RUE, field data and
40 expert knowledge, to map environmental hotspots of change and attribute them to climate and human
41 activities. One third of the area showed significant change in RUE, mainly occurring around the fall
42 of the Soviet Union (1991) or as the result of major droughts. Recent human-induced turning points in
43 ecosystem functioning were uncovered nearby Volgograd (Russia) and around Lake Balkhash
44 (Kazakhstan), attributed to recultivation, increased salinisation, and increased grazing.

45 Attribution of vegetation changes to human activity has also been done within modelling frameworks.
46 In these methods ecosystem models are used to simulate potential natural vegetation dynamics, and
47 this is compared to the observed state. The difference is attributed to human activities. Applied to the
48 Sahel region during the period of 1982–2002, it showed that people had a minor influence on

1 vegetation changes (Sequist et al., 2009). Similar model/observation comparisons performed globally
2 found that CO₂ fertilisation was the strongest forcing at global scales, with climate having regionally
3 varying effects (Mao et al., 2013; Zhu et al., 2016). Land use/land cover change was a dominant
4 forcing in localised areas. The use of this method to examine vegetation changes in China (1982–
5 2009) attributed most of the greening trend to CO₂ fertilisation and nitrogen (N) deposition (Piao et
6 al., 2015). However in some parts of northern and western China, which includes large areas of
7 drylands, Piao et al. (2015) found climate changes could be the dominant forcing. In the northern
8 extratropical land surface, the observed greening was consistent with increases in greenhouse gases
9 (notably CO₂) and the related climate change, and not consistent with a natural climate that does not
10 include anthropogenic increase in greenhouse gases (Mao et al., 2016). While many studies found
11 widespread influence of CO₂ fertilisation, it is not ubiquitous, for example, Lévesque et al. (2014)
12 found little response to CO₂ fertilisation in some tree species in Switzerland/northern Italy.



1
 2 **Figure 3.7 The Drivers of Dryland Vegetation Change.** The mean annual change in NDVI_{max} between
 3 1982 and 2015 (See Figure 3.6 for total change using Global Inventory Modelling and Mapping Studies
 4 NDVI3g v1 dataset) attributable to a) CO₂ fertilisation b) climate and c) land use. The change
 5 attributable to CO₂ fertilisation was calculated using the CO₂ fertilisation relationship described in
 6 (Franks et al., 2013). The Time Series Segmented Residual Trends (TSS-RESTREND) method (Burrell et
 7 al., 2017) applied to the CO₂ adjusted NDVI was used to separate Climate and Land Use. A multi climate
 8 dataset ensemble was used to reduce the impact of dataset errors (Burrell et al., 2018). Non-dryland
 9 regions (Aridity Index > 0.65) are masked in dark grey. Areas where the change did not meet the multi-
 10 run ensemble significance criteria, or are smaller than the error in the sensors (± 0.00001) are masked in
 11 white.

1 Using multiple extreme event attribution methodologies, Uhe et al. (2018) shows that the dominant
2 influence for droughts in Eastern Africa during October to December ‘short rains’ season is the
3 prevailing tropical SST patterns, although temperature trends mean that the current drought conditions
4 are hotter than it would have been without climate change. Similarly, Funk et al. (2019) found that
5 2017 March-June East African drought was influenced by Western Pacific SST, with high SST
6 conditions attributed to climate change.

7 There are numerous local case studies on attribution of desertification, which use different periods,
8 focus on different land uses and covers, and consider different desertification processes. For example,
9 two-thirds of the observed expansion of the Sahara Desert from 1920–2003 has been attributed to
10 natural climate cycles (the cold phase of Atlantic Multi-Decadal Oscillation and Pacific Decadal
11 Oscillation) (Thomas and Nigam, 2018). Some studies consider drought to be the main driver of
12 desertification in Africa (e.g. Masih et al., 2014). However, other studies suggest that although
13 droughts may contribute to desertification, the underlying causes are human activities (Kouba et al.,
14 2018). Brandt et al. (2016a) found that woody vegetation trends are negatively correlated with human
15 population density. Changes in land use, water pumping and flow diversion have enhanced drying of
16 wetlands and salinisation of freshwater aquifers in Israel (Inbar, 2007). The dryland territory of China
17 has been found to be very sensitive to both climatic variations and land use/land cover changes (Fu et
18 al., 2000; Liu and Tian, 2010; Zhao et al., 2013, 2006). Feng et al. (2015) shows that socioeconomic
19 factors were dominant in causing desertification in north Shanxi, China, between 1983 and 2012,
20 accounting for about 80% of desertification expansion. Successful grass establishment has been
21 impeded by overgrazing and nutrient depletion leading to the encroachment of shrubs into the
22 northern Chihuahuan Desert (USA) since the mid-1800s (Kidron and Gutschick, 2017). Human
23 activities led to rangeland degradation in Pakistan and Mongolia during 2000-2011 (Lei et al., 2011).
24 More equal shares of climatic (temperature and precipitation trends) and human factors were
25 attributed for changes in rangeland condition in China (Yang et al., 2016).

26 This kaleidoscope of local case studies demonstrates how attribution of desertification is still
27 challenging for several reasons. Firstly, desertification is caused by an interaction of different drivers
28 which vary in space and time. Secondly, in drylands, vegetation reacts closely to changes in rainfall so
29 the effect of rainfall changes on biomass needs to be ‘removed’ before attributing desertification to
30 human activities. Thirdly, human activities and climatic drivers impact vegetation/ecosystem changes
31 at different rates. Finally, desertification manifests as a gradual change in ecosystem composition and
32 structure (e.g., woody shrub invasion into grasslands). Although initiated at a limited location,
33 ecosystem change may propagate throughout an extensive area via a series of feedback mechanisms.
34 This complicates the attribution of desertification to human and climatic causes as the process can
35 develop independently once started.

36 Rasmussen et al. (2016) studied the reasons behind the overall lack of scientific agreement in trends
37 of environmental changes in the Sahel, including their causes. The study indicated that these are due
38 to differences in conceptualisations and choice of indicators, biases in study site selection, differences
39 in methods, varying measurement accuracy, differences in time and spatial scales. High resolution,
40 multi-sensor airborne platforms provide a way to address some of these issues (Asner et al., 2012).

41 The major conclusion of this section is that, with all the shortcomings of individual case studies,
42 relative roles of climatic and human drivers of desertification are location-specific and evolve over
43 time (*high confidence*). Biophysical research on attribution and socio-economic research on drivers of
44 land degradation have long studied the same topic, but in parallel, with little interdisciplinary
45 integration. Interdisciplinary work to identify typical patterns, or typologies, of such interactions of
46 biophysical and human drivers of desertification (not only of dryland vulnerability), and their relative
47 shares, done globally in comparable ways, will help in the formulation of better informed policies to
48 address desertification and achieve land degradation neutrality.

3.3. Desertification Feedbacks to Climate

While climate change can drive desertification (3.1.4.1), the process of desertification can also alter the local climate providing a feedback (Sivakumar, 2007). These feedbacks can alter the carbon cycle, and hence the level of atmospheric CO₂ and its related global climate change, or they can alter the surface energy and water budgets directly impacting the local climate. While these feedbacks occur in all climate zones (Chapter 2), here we focus on their effects in dryland regions and assess the literature concerning the major desertification feedbacks to climate. The main feedback pathways discussed throughout section 3.3 are summarised in Figure 3.8.

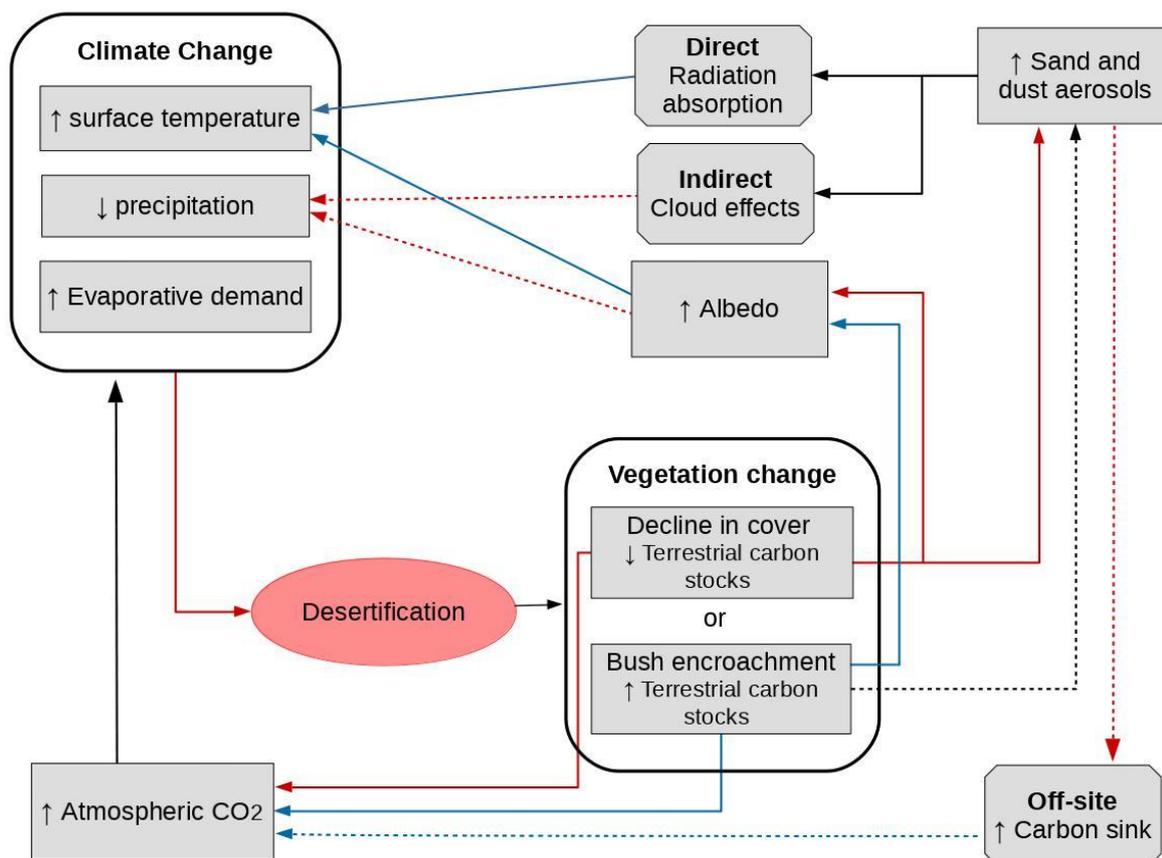
Drylands are characterised by limited soil moisture compared to humid regions. Thus, the sensible heat (heat that causes the atmospheric temperature to rise) accounts for more of the surface net radiation than latent heat (evaporation) in these regions (Wang and Dickinson, 2013). This tight coupling between the surface energy balance and the soil moisture in semi-arid and dry sub-humid zones makes these regions susceptible to land-atmosphere feedback loops that can amplify changes to the water cycle (Seneviratne et al., 2010). Changes to the land surface caused by desertification can change the surface energy budget, altering the soil moisture and triggering these feedbacks.

3.3.1. Sand and Dust Aerosols

Sand and mineral dust are frequently mobilised from sparsely vegetated drylands forming “sand storms” or “dust storms” (UNEP et al., 2016). The African continent is the most important source of desert dust, perhaps 50% of atmospheric dust comes from the Sahara (Middleton, 2017). Ginoux et al. (2012) estimated that 25% of global dust emissions have anthropogenic origins, often in drylands. These events can play an important role in the local energy balance. Through reducing vegetation cover and drying the surface conditions, desertification can increase the frequency of these events. Biological or structural soil crusts have been shown to effectively stabilise dryland soils and thus their loss, due to intense land use and/or climate change, can be expected to cause an increase in sand and dust storms (*high confidence*) (Rajot et al., 2003; Field et al., 2010; Rodriguez-Caballero et al., 2018). These sand and dust aerosols impact the regional climate in several ways (Chooari et al., 2014). The direct effect is the interception, reflection and absorption of solar radiation in the atmosphere, reducing the energy available at the land surface and increasing the temperature of the atmosphere in layers with sand and dust present (Kaufman et al., 2002; Middleton, 2017; Kok et al., 2018). The heating of the dust layer can alter the relative humidity and atmospheric stability, which can change cloud lifetimes and water content. This has been referred to as the semi-direct effect (Huang et al., 2017). Aerosols also have an indirect effect on climate through their role as cloud condensation nuclei, changing cloud radiative properties as well as the evolution and development of precipitation (Kaufman et al., 2002). While these indirect effects are more variable than the direct effects, depending on the types and amounts of aerosols present, the general tendency is toward an increase in the number, but a reduction in the size of cloud droplets, increasing the cloud reflectivity and decreasing the chances of precipitation. These effects are referred to as aerosol-radiation and aerosol-cloud interactions (Boucher et al., 2013).

There is *high confidence* that there is a negative relationship between vegetation green-up and the occurrence of dust storms (Engelstaedter et al., 2003; Fan et al., 2015; Yu et al., 2015; Zou and Zhai, 2004). Changes in groundwater can affect vegetation and the generation of atmospheric dust (Elmore et al., 2008). This can occur through groundwater processes such as the vertical movement of salt to the surface causing salinisation, supply of near surface soil moisture, and sustenance of groundwater dependent vegetation. Groundwater dependent ecosystems have been identified in many dryland regions around the world (Decker et al., 2013; Lamontagne et al., 2005; Patten et al., 2008). In these locations declining groundwater levels can decrease vegetation cover. Cook et al., (2009) found that dust aerosols intensified the “dust bowl” drought in North America during the 1930s.

1 By decreasing the amount of green cover and hence increasing the occurrence of sand and dust
 2 storms, desertification will increase the amount of shortwave cooling associated with the direct effect
 3 (*high confidence*). There is *medium confidence* that the semi-direct and indirect effects of this dust
 4 would tend to decrease precipitation and hence provide a positive feedback to desertification (Huang
 5 et al., 2009; Konare et al., 2008; Rosenfeld et al., 2001; Solomon et al., 2012; Zhao et al., 2015).
 6 However, the combined effect of dust has also been found to increase precipitation in some areas
 7 (Islam and Almazroui, 2012; Lau et al., 2009; Sun et al., 2012). The overall combined effect of dust
 8 aerosols on desertification remains uncertain with *low agreement* between studies that find positive
 9 (Huang et al., 2014), negative (Miller et al., 2004) or no feedback on desertification (Zhao et al.,
 10 2015).



11
 12 **Figure 3.8 Schematic of main pathways through which desertification can feedback on climate as**
 13 **discussed in section 3.4. Note: red arrows indicate a positive effect. Blue arrows indicate a negative effect.**
 14 **Black arrows indicate an indeterminate effect (potentially both positive and negative). Solid arrows are**
 15 **direct while dashed arrows are indirect.**

16 3.3.1.1. Off-site Feedbacks

17 Aerosols can act as a vehicle for the long-range transport of nutrients to oceans (Jickells et al., 2005;
 18 Okin et al., 2011) and terrestrial land surfaces (Das et al., 2013). In several locations, notably the
 19 Atlantic Ocean, west of northern Africa and the Pacific Ocean east of northern China, a considerable
 20 amount of mineral dust aerosols, sourced from nearby drylands, reaches the oceans. It was estimated
 21 that 60% of dust transported off Africa is deposited in the Atlantic Ocean (Kaufman et al., 2005),
 22 while 50% of the dust generated in Asia reaches the Pacific Ocean or further (Uno et al., 2009; Zhang
 23 et al., 1997). The Sahara is also a major source of dust for the Mediterranean basin (Varga et al.,
 24 2014). The direct effect of atmospheric dust over the ocean was found to be a cooling of the ocean
 25 surface (*limited evidence, high agreement*) (Evan and Mukhopadhyay, 2010; Evan et al., 2009) with
 26 the tropical North Atlantic mixed layer cooling by over 1°C (Evan et al., 2009).

1 It has been suggested that dust may act as a source of nutrients for the upper ocean biota, enhancing
2 the biological activity and related C sink (*medium evidence, low agreement*) (Lenes et al., 2001; Shaw
3 et al., 2008; Neuer et al., 2004). The overall response depends on the environmental controls on the
4 ocean biota, the type of aerosols including their chemical constituents, and the chemical environment
5 in which they dissolve (Boyd et al., 2010).

6 Dust deposited on snow can increase the amount of absorbed solar radiation leading to more rapid
7 melting (Painter et al., 2018), impacting a region's hydrological cycle (*high confidence*). Dust
8 deposition on snow and ice has been found in many regions of the globe (e.g. Painter et al., 2018;
9 Kaspari et al., 2014; Qian et al., 2015; Painter et al. 2013), however quantification of the effect
10 globally and estimation of future changes in the extent of this effect remain knowledge gaps.

11 **3.3.2. Changes in Surface Albedo**

12 Increasing surface albedo in dryland regions will impact the local climate, decreasing surface
13 temperature and precipitation, and provide a positive feedback on the albedo (*high confidence*)
14 (Charney et al., 1975). This albedo feedback can occur in desert regions worldwide (Zeng and Yoon,
15 2009). Similar albedo feedbacks have also been found in regional studies over the Middle East
16 (Zaitchik et al., 2007), Australia (Evans et al., 2017; Meng et al., 2014a,b), South America (Lee and
17 Berbery, 2012) and the USA (Zaitchik et al., 2013).

18 Recent work has also found albedo in dryland regions can be associated with soil surface communities
19 of lichens, mosses and cyanobacteria (Rodriguez-Caballero et al., 2018). These communities compose
20 the soil crust in these ecosystems and due to the sparse vegetation cover, directly influence the albedo.
21 These communities are sensitive to climate changes with field experiments indicating albedo changes
22 greater than 30% are possible. Thus, changes in these communities could trigger surface albedo
23 feedback processes (*limited evidence, high agreement*) (Rutherford et al., 2017).

24 A further pertinent feedback relationship exists between changes in land-cover, albedo, C stocks and
25 associated GHG emissions, particularly in drylands with low levels of cloud cover. One of the first
26 studies to focus on the subject was Rotenberg and Yakir (2010), who used the concept of 'radiative
27 forcing' to compare the relative climatic effect of a change in albedo with a change in atmospheric
28 GHGs due to the presence of forest within drylands. Based on this analysis, it was estimated that the
29 change in surface albedo due to the degradation of semi-arid areas has decreased radiative forcing in
30 these areas by an amount equivalent to approximately 20% of global anthropogenic GHG emissions
31 between 1970 and 2005 (Rotenberg and Yakir, 2010).

32 **3.3.3. Changes in Vegetation and Greenhouse Gas Fluxes**

33 Terrestrial ecosystems have the ability to alter atmospheric GHGs through a number of processes
34 (Schlesinger et al., 1990). This may be through a change in plant and soil C stocks, either sequestering
35 atmospheric carbon dioxide (CO₂) during growth or releasing C during combustion and respiration, or
36 through processes such as enteric fermentation of domestic and wild ruminants that lead to the release
37 of methane and nitrous oxide (Sivakumar, 2007). It is estimated that 241-470 Gt C is stored in dryland
38 soils (top 1m Lal, 2004; Plaza et al., 2018). When evaluating the effect of desertification, the net
39 balance of all the processes and associated GHG fluxes needs to be considered.

40 Desertification usually leads to a loss in productivity and a decline in above- and below-ground C
41 stocks (Abril et al., 2005; Asner et al., 2003). Drivers such as overgrazing lead to a decrease in both
42 plant and SOC pools (Abdalla et al., 2018). While dryland ecosystems are often characterised by open
43 vegetation, not all drylands have low biomass and C stocks in an intact state (Lechmere-Oertel et al.,
44 2005; Maestre et al., 2012). Vegetation types such as the subtropical thicket of South Africa have over
45 70 t C ha⁻¹ in an intact state, greater than 60% of which is released into the atmosphere during
46 degradation through overgrazing (Lechmere-Oertel et al., 2005; Powell, 2009). In comparison, semi-

1 arid grasslands and savannas with similar rainfall, may have only 5-35 t C ha⁻¹ (Scholes and Walker,
2 1993; Woomeer et al., 2004)

3 At the same time, it is expected that a decline in plant productivity may lead to a decrease in fuel
4 loads and a reduction in CO₂, nitrous oxide and methane emissions from fire. In a similar manner,
5 decreasing productivity may lead to a reduction in ruminant animals that in turn would decrease
6 methane emissions. Few studies have focussed on changes in these sources of emissions due to
7 desertification and it remains a field that requires further research.

8 In comparison to desertification through the suppression of primary production, the process of woody
9 plant encroachment can result in significantly different climatic feedbacks. Increasing woody plant
10 cover in open rangeland ecosystems leads to an increase in woody C stocks both above- and below-
11 ground (Asner et al., 2003; Hughes et al., 2006; Petrie et al., 2015; Li et al., 2016). Within the
12 drylands of Texas, shrub encroachment led to a 32% increase in aboveground C stocks over a period
13 of 69 years (3.8 t C ha⁻¹ to 5.0 t C ha⁻¹) (Asner et al., 2003). Encroachment by taller woody species,
14 can lead to significantly higher observed biomass and C stocks, for example, encroachment by
15 *Dichrostachys cinerea* and several *Vachellia* species in the sub-humid savannas of north-west South
16 Africa led to an increase of 31–46 t C ha⁻¹ over a 50–65 year period (1936–2001) (Hudak et al., 2003).
17 In terms of potential changes in SOC stocks, the effect may be dependent on annual rainfall and soil
18 type. Woody cover generally leads to an increase in SOC stocks in drylands that have less than 800
19 mm of annual rainfall, while encroachment can lead to a loss of soil C in more humid ecosystems
20 (Barger et al., 2011; Jackson et al., 2002).

21 The suppression of the grass layer through the process of woody encroachment may lead to a decrease
22 in C stocks within this relatively small C pool (Magandana, 2016). Conversely, increasing woody
23 cover may lead to a decrease and even halt in surface fires and associated GHG emissions. In analysis
24 of drivers of fire in southern Africa, Archibald et al. (2009) note that there is a potential threshold
25 around 40% canopy cover, above which surface grass fires are rare. Whereas there have been a
26 number of studies on changes in C stocks due to desertification in North America, southern Africa and
27 Australia, a global assessment of the net change in C stocks as well as fire and ruminant GHG
28 emissions due to woody plant encroachment has not been done yet.

29

30 **3.4. Desertification Impacts on Natural and Socio-Economic Systems** 31 **under Climate Change**

32 **3.4.1. Impacts on Natural and Managed Ecosystems**

33 **3.4.1.1. Impacts on Ecosystems and their Services in Drylands**

34 The Millenium Ecosystem Assesment (2005) proposed four classes of ecosystem services:
35 provisioning, regulating, supporting and cultural services (Cross-Chapter Box 8: Ecosystem Services,
36 Chapter 6). These ecosystem services in drylands are vulnerable to the impacts of climate change due
37 to high variability in temperature, precipitation and soil fertility (Enfors and Gordon, 2008;
38 Mortimore, 2005). There is *high confidence* that desertification processes such as soil erosion,
39 secondary salinisation, and overgrazing have negatively impacted provisioning ecosystem services in
40 drylands, particularly food and fodder production (Majeed and Muhammad, 2019; Mirzabaev et al.,
41 2016; Qadir et al., 2009; Van Loo et al., 2017; Tokbergenova et al., 2018) (3.4.2.2). Zika and Erb
42 (2009) reported an estimation of NPP losses between 0.8 and 2.0 Gt C yr⁻¹ due to desertification,
43 comparing the potential NPP and the NPP calculated for the year 2000. In terms of climatic factors,
44 although climatic changes between 1976 and 2016 were found overall favourable for crop yields in
45 Russia (Ivanov et al., 2018), yield decreases of up to 40-60% in dryland areas were caused by severe
46 and extensive droughts (Ivanov et al., 2018). Increase in temperature can have a direct impact on

1 animals in the form of increased physiological stress (Rojas-Downing et al., 2017), increased water
2 requirements for drinking and cooling, a decrease in the production of milk, meat and eggs, increased
3 stress during conception and reproduction (Nardone et al., 2010) or an increase in seasonal diseases
4 and epidemics (Thornton et al., 2009; Nardone et al., 2010). Furthermore, changes in temperature can
5 indirectly impact livestock through reducing the productivity and quality of feed crops and forages
6 (Thornton et al., 2009; Polley et al., 2013). On the other hand, fewer days with extreme cold
7 temperatures during winter in the temperate zones are associated with lower livestock mortality. The
8 future projection of impacts on ecosystems is presented in section 3.5.2.

9 Over-extraction is leading to groundwater depletion in many dryland areas (*high confidence*) (Mudd,
10 2000; Mays, 2013; Mahmud and Watanabe, 2014; Jolly et al., 2008). Globally, groundwater reserves
11 have been reduced since 1900, with the highest rate of estimated reductions of 145 km³ yr⁻¹ between
12 2000 and 2008 (Konikow, 2011). Some arid lands are very vulnerable to groundwater reductions,
13 because the current natural recharge rates are lower than during the previous wetter periods (e.g.,
14 Atacama Desert and Nubian aquifer system in Africa; (Squeo et al., 2006; Mahmud and Watanabe,
15 2014; Herrera et al., 2018).

16 Among regulating services, desertification can influence levels of atmospheric CO₂. In drylands, the
17 majority of C is stored below ground in the form of biomass and SOC (FAO, 1995) (3.3.3). Land-use
18 changes often lead to reductions in SOC and organic matter inputs into soil (Albaladejo et al., 2013;
19 Almagro et al., 2010; Hoffmann et al., 2012; Lavee et al., 1998; Rey et al., 2011), increasing soil
20 salinity and soil erosion (Lavee et al., 1998; Martinez-Mena et al., 2008). In addition to the loss of
21 soil, erosion reduces soil nutrients and organic matter, thereby impacting land's productive capacity.
22 To illustrate, soil erosion by water is estimated to result in the loss of 23–42 M t of N and 14.6–26.4
23 Mt of phosphorus from soils annually in the world (Pierzynski et al., 2017).

24 Precipitation, by affecting soil moisture content, is considered to be the principal determinant of the
25 capacity of drylands to sequester C (Fay et al., 2008; Hao et al., 2008; Mi et al., 2015; Serrano-Ortiz
26 et al., 2015; Vargas et al., 2012; Sharkhuu et al., 2016). Lower annual rainfall resulted in the release
27 of C into the atmosphere for a number of sites located in Mongolia, China and North America
28 (Biederman et al., 2017; Chen et al., 2009; Fay et al., 2008; Hao et al., 2008; Mi et al., 2015;
29 Sharkhuu et al., 2016). Low soil water availability promotes soil microbial respiration, yet there is
30 insufficient moisture to stimulate plant productivity (Austin et al., 2004), resulting in net C emissions
31 at an ecosystem level. Under even drier conditions, photodegradation of vegetation biomass may often
32 constitute an additional loss of C from ecosystem (Rutledge et al., 2010). In contrast, years of good
33 rainfall in drylands resulted in the sequestration of C (Biederman et al., 2017; Chen et al., 2009; Hao
34 et al., 2008) In an exceptionally rainy year (2011) in the southern hemisphere, the semiarid
35 ecosystems of this region contributed 51% of the global net C sink (Poulter et al., 2014). These results
36 suggest that arid ecosystems could be an important global C sink depending on soil water availability
37 (*medium evidence, high agreement*). However, drylands are generally predicted to become warmer
38 with an increasing frequency of extreme drought and high rainfall events (Donat et al., 2016).

39 When desertification reduces vegetation cover, this alters the soil surface, affecting the albedo and the
40 water balance (Gonzalez-Martin et al., 2014) (3.3). In such situations, erosive winds have no more
41 obstacles, which favour the occurrence of wind erosion and dust storms. Mineral aerosols have an
42 important influence on the dispersal of soil nutrients and lead to changes in soil characteristics
43 (Goudie and Middleton, 2001; Middleton, 2017). Thereby, the soil formation as a supporting
44 ecosystem service is negatively affected (3.3.1.). Soil erosion by wind results in a loss of fine soil
45 particles (silt and clay), reducing the ability of soil to sequester C (Wiesmeier et al., 2015). Moreover,
46 dust storms reduce crop yields by loss of plant tissue caused by sandblasting (resulting loss of plant
47 leaves and hence reduced photosynthetic activity (Field et al., 2010), exposing crop roots, crop seed
48 burial under sand deposits, and leading to losses of nutrients and fertiliser from top soil (Stefanski and

1 Sivakumar, 2009). Dust storms also impact crop yields by reducing the quantity of water available for
2 irrigation because it could decrease the storage capacity of reservoirs by siltation and block
3 conveyance canals (Middleton, 2017; Middleton and Kang, 2017; Stefanski and Sivakumar, 2009).
4 Livestock productivity is reduced by injuries caused by dust storms (Stefanski and Sivakumar, 2009).
5 Additionally, dust storms favor the dispersion of microbial and plants species, which can make local
6 endemic species vulnerable to extinction and promote the invasion of plant and microbial species
7 (Asem and Roy, 2010; Womack et al., 2010). Dust storms increase microbial species in remote sites
8 (*high confidence*); (Kellogg et al., 2004; Prospero et al., 2005; Griffin et al., 2006; Schlesinger et al.,
9 2006; Griffin, 2007; De Deckker et al., 2008; Jeon et al., 2011; Abed et al., 2012; Favet et al., 2013;
10 Woo et al., 2013; Pointing and Belnap, 2014).

11

12 **3.4.1.2. Impacts on Biodiversity: Plant and Wildlife**

13 **3.4.1.2.1. Plant Biodiversity**

14 Over 20% of global plant biodiversity centres are located within drylands (White and Nackoney,
15 2003). Plant species located within these areas are characterised by high genetic diversity within
16 populations (Martínez-Palacios et al., 1999). The plant species within these ecosystems are often
17 highly threatened by climate change and desertification (Millennium Ecosystem Assessment, 2005b;
18 Maestre et al., 2012). Increasing aridity exacerbates the risk of extinction of some plant species,
19 especially those that are already threatened due to small populations or restricted habitats (Gitay et al.,
20 2002). Desertification, including through land use change, already contributed to the loss of
21 biodiversity across drylands (*medium confidence*) (Newbold et al., 2015; Wilting et al., 2017). For
22 example, species richness decreased from 234 species in 1978 to 95 in 2011 following long periods of
23 drought and human driven degradation on the steppe land of south western Algeria (Observatoire du
24 Sahara et du Sahel, 2013). Similarly, drought and overgrazing led to loss of biodiversity in Pakistan,
25 where only drought-adapted species have by now survived on arid rangelands (Akhter and Arshad,
26 2006). Similar trends were observed in desert steppes of Mongolia (Khishigbayar et al., 2015). In
27 contrast, the increase in annual moistening of southern European Russia from the late 1980s to the
28 beginning of the 21st century caused the restoration of steppe vegetation, even under conditions of
29 strong anthropogenic pressure (Ivanov et al., 2018). The seed banks of annual species can often
30 survive over the long-term, germinating in wet years, suggesting that these species could be resilient
31 to some aspects of climate change (Vetter et al., 2005). Yet, Hiernaux and Houérou (2006) showed
32 that overgrazing in the Sahel tended to decrease the seed bank of annuals which could make them
33 vulnerable to climate change over time. Perennial species, considered as the structuring element of the
34 ecosystem, are usually less affected as they have deeper roots, xeromorphic properties and
35 physiological mechanisms that increase drought tolerance (Le Houérou, 1996). However, in North
36 Africa, long-term monitoring (1978–2014) has shown that important plant perennial species have also
37 disappeared due to drought (*Stipa tenacissima* and *Artemisia herba alba*) (Hirche et al., 2018;
38 Observatoire du Sahara et du Sahel, 2013). The aridisation of the climate in the south of Eastern
39 Siberia led to the advance of the steppes to the north and to the corresponding migration of steppe
40 mammal species between 1976 and 2016 (Ivanov et al., 2018). The future projection of impacts on
41 plant biodiversity is presented in the section 3.5.2.

42 **3.4.1.2.2. Wildlife biodiversity**

43 Dryland ecosystems have high levels of faunal diversity and endemism (MEA, 2005; Whitford,
44 2002). Over 30% of the endemic bird areas are located within these regions, which is also home to
45 25% of vertebrate species (Maestre et al., 2012; MEA, 2005). Yet, many species within drylands are
46 threatened with extinction (Durant et al., 2014; Walther, 2016). Habitat degradation and
47 desertification are generally associated with biodiversity loss (Ceballos et al. 2010; Tang et al. 2018;
48 Newbold et al. 2015). The “grazing value” of land declines with both a reduction in vegetation cover
49 and shrub encroachment, with the former being more detrimental to native vertebrates (Parsons et al.,

1 2017). Conversely, shrub encroachment may buffer desertification by increasing resource and
2 microclimate availability, resulting in an increase in vertebrate species abundance and richness
3 observed in the shrub encroached arid grasslands of North America (Whitford, 1997) and Australia
4 (Parsons et al., 2017). However, compared to historically resilient drylands, these encroached habitats
5 and their new species assemblages may be more sensitive to droughts, which may more prevalent
6 with climate change (Schooley et al., 2018). Mammals and birds may be particularly sensitive to
7 droughts because they rely on evaporative cooling to maintain their body temperatures within an
8 optimal range (Hetem et al., 2016) and risk lethal dehydration in water limited environments (Albright
9 et al., 2017). The direct effects of reduced rainfall and water availability are *likely* to be exacerbated
10 by the indirect effects of desertification through a reduction in primary productivity. A reduction in
11 the quality and quantity of resources available to herbivores due to desertification under changing
12 climate can have knock-on consequences for predators and may ultimately disrupt trophic cascades
13 (*limited evidence, low agreement*) (Rey et al. 2017; Walther 2010). Reduced resource availability may
14 also compromise immune response to novel pathogens, with increased pathogen dispersal associated
15 with dust storms (Zinabu et al., 2018). Responses to desertification are species-specific and
16 mechanistic models are not yet able to accurately predict individual species responses to the many
17 factors associated with desertification (Fuller et al., 2016).

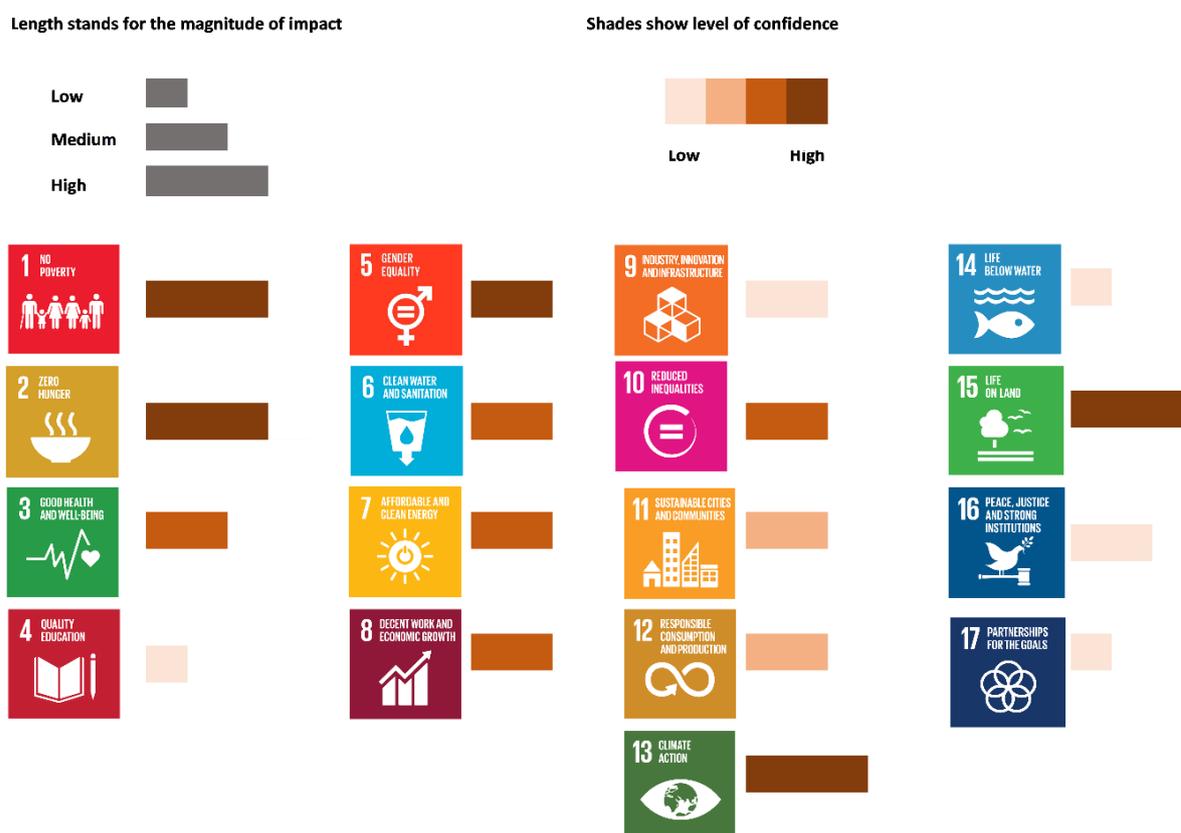
18 19 **3.4.2. Impacts on Socio-economic Systems**

20 Combined impacts of desertification and climate change on socio-economic development in drylands
21 are complex. Figure 3.9 schematically represents our qualitative assessment of the magnitudes and the
22 uncertainties associated with these impacts on attainment of the SDGs in dryland areas (UN, 2015).
23 The impacts of desertification and climate change are difficult to isolate from the effects of other
24 socio-economic, institutional and political factors (Pradhan et al., 2017). However, there is *high*
25 *confidence* that climate change will exacerbate the vulnerability of dryland populations to
26 desertification, and that the combination of pressures coming from climate change and desertification
27 will diminish opportunities for reducing poverty, enhancing food and nutritional security, empowering
28 women, reducing disease burden, improving access to water and sanitation. Desertification is
29 embedded in SDG 15 (target 15.3) and climate change is under SDG 13, the *high confidence* and high
30 magnitude impacts depicted for these SDGs (Figure 3.9) indicate that the interactions between
31 desertification and climate change strongly affect the achievement of the targets of SDGs 13 and 15.3,
32 pointing at the need for the coordination of policy actions on land degradation neutrality and
33 mitigation and adaptation to climate change. The following subsections present the literature and the
34 assessment which serve as the basis for Figure 3.9.

35 **3.4.2.1 Impacts on Poverty**

36 Climate change has a high potential to contribute to poverty particularly through the risks coming
37 from extreme weather events (Olsson et al., 2014). However, the evidence rigorously attributing
38 changes in observed poverty to climate change impacts is currently not available. On the other hand,
39 most of the research on links between poverty and desertification (or more broadly, land degradation)
40 focused on whether or not poverty is a cause of land degradation (Gerber et al., 2014; Vu et al., 2014;
41 Way, 2016; 4.7.1). The literature measuring to what extent desertification contributed to poverty
42 globally is lacking: the related literature remains qualitative or correlational (Barbier and Hochard,
43 2016). At the local level, on the other hand, there is *limited evidence* and *high agreement* that
44 desertification increased multidimensional poverty. For example, Diao and Sarpong (2011) estimated
45 that land degradation lowered agricultural incomes in Ghana by USD 4.2 billion between 2006 and
46 2015, increasing the national poverty rate by 5.4% in 2015. Land degradation increased the
47 probability of households becoming poor by 35% in Malawi and 48% in Tanzania (Kirui, 2016).
48 Desertification in China was found to have resulted in substantial losses in income, food production

1 and jobs (Jiang et al., 2014). On the other hand, Ge et al. (2015) indicated that desertification was
 2 positively associated with growing incomes in Inner Mongolia in China in the short run since no costs
 3 were incurred for SLM, while in the long run higher incomes allowed allocation of more investments
 4 to reduce desertification. This relationship corresponds to the Environmental Kuznets Curve, which
 5 posits that environmental degradation initially rises and subsequently falls with rising income (e.g.
 6 Stern, 2017). There is *limited evidence* on the validity of this hypothesis regarding desertification.



9
10 **Figure 3.9 Socio-economic impacts of desertification and climate change with the SDG framework**

11 3.4.2.2 Impacts on Food and Nutritional Insecurity

12 About 821 million people globally were food insecure in 2017, of whom 63% in Asia, 31% in Africa
 13 and 5% in Latin America and the Caribbean (FAO et al., 2018). The global number of food insecure
 14 people rose by 37 million since 2014. Changing climate variability, combined with a lack of climate
 15 resilience, was suggested as a key driver of this increase (FAO et al., 2018). Sub-Saharan Africa, East
 16 Africa and South Asia had the highest share of undernourished populations in the world in 2017, with
 17 28.8%, 31.4% and 33.7%, respectively (FAO et al., 2018). The major mechanism through which
 18 climate change and desertification affect food security is through their impacts on agricultural
 19 productivity. There is *robust evidence* pointing to negative impacts of climate change on crop yields
 20 in dryland areas (*high agreement*) (Hochman et al., 2017; Nelson et al., 2010; Zhao et al., 2017; 3.4.1;
 21 5.2.2; 4.7.2). There is also *robust evidence and high agreement* on the losses in agricultural
 22 productivity and incomes due to desertification (Kirui, 2016; Moussa et al., 2016; Mythili and
 23 Goedecke, 2016; Tun et al., 2015). Nkonya et al. (2016a) estimated that cultivating wheat, maize, and
 24 rice with unsustainable land management practices is currently resulting in global losses of USD 56.6
 25 billion annually, with another USD 8.7 billion of annual losses due to lower livestock productivity
 26 caused by rangeland degradation. However, the extent to which these losses affected food insecurity
 27 in dryland areas is not known. Lower crop yields and higher agricultural prices worsen existing food

1 insecurity, especially for net-food buying rural households and urban dwellers. Climate change and
2 desertification are not the sole drivers of food insecurity, but especially in the areas with high
3 dependence on agriculture, they are among the main contributors.

4 **3.4.2.3 Impacts on Human Health through Dust Storms**

5 The frequency and intensity of dust storms are increasing due to land use and land cover changes and
6 climate-related factors (2.4) particularly in some regions of the world such as the Arabian Peninsula
7 (Jish Prakash et al., 2015; Yu et al., 2015; Gherboudj et al., 2017; Notaro et al., 2013; Yu et al. 2013;
8 Alobaidi et al., 2017; Maghrabi et al., 2011; Almazroui et al. 2018) and broader Middle East (Rashki
9 et al., 2012; Türkeş, 2017; Namdari et al., 2018) as well as Central Asia (Indoitu et al., 2015; Xi and
10 Sokolik, 2015), with growing negative impacts on human health (Díaz et al., 2017; Goudarzi et al.,
11 2017; Goudie, 2014; Samoli et al., 2011) (*high confidence*). Dust storms transport particulate matter,
12 pollutants, pathogens and potential allergens that are dangerous for human health over long distances
13 (Goudie and Middleton, 2006; Sprigg, 2016). Particulate matter (PM), i.e. the suspended particles in
14 the air having sizes of 10 micrometre (PM10) or less, have damaging effects on human health (Díaz et
15 al., 2017; Goudarzi et al., 2017; Goudie, 2014; Samoli et al., 2011). The health effects of dust storms
16 are largest in areas in the immediate vicinity of their origin, primarily the Sahara Desert, followed by
17 Central and Eastern Asia, the Middle East and Australia (Zhang et al., 2016), however, there is *robust*
18 *evidence* showing that the negative health effects of dust storms reach a much wider area (Bennett et
19 al., 2006; Díaz et al., 2017; Kashima et al., 2016; Lee et al., 2014; Samoli et al., 2011; Zhang et al.,
20 2016). The primary health effects of dust storms include damage to the respiratory and cardiovascular
21 systems (Goudie, 2013). Dust particles with a diameter smaller than 2.5µm were associated with
22 global cardiopulmonary mortality of about 402,000 people in 2005, with 3.47 million years of life lost
23 in that single year (Giannadaki et al., 2014). If globally only 1.8% of cardiopulmonary deaths were
24 caused by dust storms, in the countries of the Sahara region, Middle East, South and East Asia, dust
25 storms were suggested to be the reason for 15–50% of all cardiopulmonary deaths (Giannadaki et al.,
26 2014). A 10µgm⁻³ increase in PM10 dust particles was associated with mean increases in non-
27 accidental mortality from 0.33% to 0.51% across different calendar seasons in China, Japan and South
28 Korea (Kim et al., 2017). The percentage of all-cause deaths attributed to fine particulate matter in
29 Iranian cities affected by Middle Eastern dust storms (MED) were 0.56–5.02%, while the same
30 percentage for non-affected cities were 0.16–4.13% (Hopke et al., 2018). The Meningococcal
31 Meningitis epidemics occur in the Sahelian region during the dry seasons with dusty conditions
32 (Agier et al., 2012; Molesworth et al., 2003). Despite a strong concentration of dust storms in the
33 Sahel, North Africa, the Middle East and Central Asia, there is relatively little research on human
34 health impacts of dust storms in these regions. More research on health impacts and related costs of
35 dust storms as well as on public health response measures can help in mitigating these health impacts.

36 **3.4.2.4. Impacts on Gender Equality**

37 Environmental issues such as desertification and impacts of climate change have been increasingly
38 investigated through a gender lens (Bose; Broeckhoven and Cliquet, 2015; Kaijser and Kronsell,
39 2014; Kiptot et al., 2014; Villamor and van Noordwijk, 2016). There is *medium evidence* and *high*
40 *agreement* that women will be impacted more than men by environmental degradation (Arora-
41 Jonsson, 2011; Gurung et al., 2006; Cross-Chapter Box 11: Gender, Chapter 7). Socially structured
42 gender-specific roles and responsibilities, daily activities, access and control over resources, decision-
43 making and opportunities lead men and women to interact differently with natural resources and
44 landscapes. For example, water scarcity affected women more than men in rural Ghana as they had to
45 spend more time in fetching water, which has implications on time allocations for other activities
46 (Ahmed et al., 2016). Despite the evidence pointing to differentiated impact of environmental
47 degradation on women and men, gender issues have been marginally addressed in many land
48 restoration and rehabilitation efforts, which often remain gender-blind. Although there is *robust*
49

1 *evidence* on the location-specific impacts of climate change and desertification on gender equality,
2 however, there is *limited evidence* on the gender-related impacts of land restoration and rehabilitation
3 activities. Women are usually excluded from local decision making on actions regarding
4 desertification and climate change. Socially constructed gender-specific roles and responsibilities are
5 not static because they are shaped by other factors such as wealth, age, ethnicity, and formal education
6 (Kajiser and Kronsell, 2014; Villamor et al., 2014). Hence, women's and men's environmental
7 knowledge and priorities for restoration often differ (Sijapati Basnett et al., 2017). In some areas
8 where sustainable land options (e.g. agroforestry) are being promoted, women were not able to
9 participate due to culturally-embedded asymmetries in power relations between men and women
10 (Catacutan and Villamor, 2016). Nonetheless, women particularly in the rural areas remain heavily
11 involved in securing food for their households. Food security for them is associated with land
12 productivity and women's contribution to address desertification is crucial.

13 **3.4.2.5. Impacts on Water Scarcity and Use**

14 Reduced water retention capacity of degraded soils amplifies floods (de la Paix et al., 2011),
15 reinforces degradation processes through soil erosion, and reduces annual intake of water to aquifers,
16 exacerbating existing water scarcities (Le Roux et al., 2017; Cano et al., 2018). Reduced vegetation
17 cover and more intense dust storms were found to intensify droughts (Cook et al., 2009). Moreover,
18 secondary salinisation in the irrigated drylands often requires leaching with considerable amounts of
19 water (Greene et al., 2016; Wichelns and Qadir, 2015). Thus, different types of soil degradation
20 increase water scarcity both through lower water quantity and quality (Liu et al., 2017; Liu et al.,
21 2016c). All these processes reduce water availability for other needs. In this context, climate change
22 will further intensify water scarcity in some dryland areas and increase the frequency of droughts
23 (*medium confidence*) (2.2; IPCC, 2013; Zheng et al., 2018). Higher water scarcity may imply growing
24 use of wastewater effluents for irrigation (Pedrero et al., 2010). The use of untreated wastewater
25 exacerbates soil degradation processes (Tal, 2016; Singh et al., 2004; Qishlaqi et al., 2008; Hanjra et
26 al., 2012), in addition to negative human health impacts (Faour-Klingbeil and Todd, 2018; Hanjra et
27 al., 2012). Climate change, thus, will amplify the need for integrated land and water management for
28 sustainable development.

29 **3.4.2.6 Impacts on Energy Infrastructure through Dust Storms**

30 Desertification leads to conditions that favour the production of dust storms (*high confidence*) (3.3.1).
31 There is *robust evidence and high agreement* that dust storms negatively affect the operational
32 potential of solar and wind power harvesting equipment through dust deposition, reduced reach of
33 solar radiation and increasing blade surface roughness, and can also reduce effective electricity
34 distribution in high-voltage transmission lines (Zidane et al., 2016; Costa et al., 2016; Lopez-Garcia et
35 al., 2016; Maliszewski et al., 2012; Mani and Pillai, 2010; Mejia and Kleissl, 2013; Mejia et al., 2014;
36 Middleton, 2017; Sarver et al., 2013; Kaufman et al., 2002; Kok et al., 2018). Direct exposure to
37 desert dust storm can reduce energy generation efficiency of solar panels by 70–80% in one hour
38 (Ghazi et al., 2014). Saidan et al.(2016) indicated that in the conditions of Baghdad, Iraq, one month
39 exposure to weather reduced the efficiency of solar modules by 18.74% due to dust deposition. In
40 Atacama desert, Chile, one month exposure reduced thin-film solar module performance by 3.7-4.8%
41 (Fuentelba et al., 2015). This has important implications for climate change mitigation efforts using
42 the expansion of solar and wind energy generation in dryland areas for substituting fossil fuels.
43 Abundant access to solar energy in many dryland areas makes them high potential locations for the
44 installation of solar energy generating infrastructure. Increasing desertification, resulting in higher
45 frequency and intensity of dust storms imposes additional costs for climate change mitigation through
46 deployment of solar and wind energy harvesting facilities in dryland areas. Most frequently used
47 solutions to this problem involve physically wiping or washing the surface of solar devices with
48 water. These result in additional costs and excessive use of already scarce water resources and labour

1 (Middleton, 2017). The use of special coatings on the surface of solar panels can help prevent the
2 deposition of dusts (Costa et al., 2016; Costa et al., 2018; Gholami et al., 2017).

3 **3.4.2.7 Impacts on Transport Infrastructure through Dust Storms and Sand Movement**

4 Dust storms and movement of sand dunes often threaten the safety and operation of railway and road
5 infrastructure in arid and hyper-arid areas, and can lead to road and airport closures due to reductions
6 in visibility. For example, the dust storm on 10th March 2009 over Riyadh was assessed to be the
7 strongest in the previous two decades in Saudi Arabia, causing limited visibility, airport shutdown and
8 damages to infrastructure and environment across the city (Maghrabi et al., 2011). There are
9 numerous historical examples of how moving sand dunes led to the forced decommissioning of early
10 railway lines built in Sudan, Algeria, Namibia and Saudi Arabia in the late 19th and early 20th century
11 (Bruno et al., 2018). Currently, the highest concentration of railways vulnerable to sand movements
12 are located in north-western China, Middle East and North Africa (Bruno et al., 2018; Cheng and
13 Xue, 2014). In China, sand dune movements are periodically disrupting the railway transport in
14 Linhai-Ceke line in north-western China and Lanzhou-Xinjiang High-speed Railway in western
15 China, with considerable clean-up and maintenance costs (Bruno et al., 2018; Zhang et al., 2010).
16 There are large-scale plans for expansion of railway networks in arid areas of China, Central Asia,
17 North Africa, the Middle East, and Eastern Africa. For example, “The Belt and Road Initiative”
18 promoted by China, the Gulf Railway project by the countries of the Arab Gulf Cooperation Council
19 (GCC), or Lamu Port-South Sudan- Ethiopia Transport Corridor in Eastern Africa. These investments
20 have long-term return and operation periods. Their construction and associated engineering solutions
21 will therefore benefit from careful consideration of potential desertification and climate change effects
22 on sand storms and dune movements.

23 **3.4.2.8 Impacts on Conflicts**

24 There is *low confidence* in climate change and desertification leading to violent conflicts. There is
25 *medium evidence* and *low agreement* that climate change and desertification contribute to already
26 existing conflict potentials (Herrero, 2006; von Uexkull et al., 2016; Theisen, 2017; Olsson, 2017;
27 Wischnath and Buhaug, 2014; 4.7.3). To illustrate, Hsiang et al. (2013) found that each one standard
28 deviation increase in temperature or rainfall was found to increase interpersonal violence by 4% and
29 intergroup conflict by 14% (Hsiang et al., 2013). However, this conclusion was disputed by Buhaug et
30 al., (2014), who found no evidence linking climate variability to violent conflict after replicating
31 Hsiang et al. (2013) by studying only violent conflicts. Almer et al. (2017) found that a one-standard
32 deviation increase in dryness raised the likelihood of riots in Sub-Saharan African countries by 8.3%
33 during the 1990–2011 period. On the other hand, Owain and Maslin (2018) found that droughts and
34 heatwaves were not significantly affecting the level of regional conflict in East Africa. Similarly, it
35 was suggested that droughts and desertification in the Sahel have played a relatively minor role in the
36 conflicts in the Sahel in the 1980s, with the major reasons for the conflicts during this period being
37 political, especially the marginalisation of pastoralists (Benjaminsen, 2016), corruption and rent-
38 seeking (Benjaminsen et al., 2012). Moreover, the role of environmental factors as the key drivers of
39 conflicts were questioned in the case of Sudan (Verhoeven, 2011) and Syria (De Châtel, 2014).
40 Selection bias, when the literature focuses on the same few regions where conflicts occurred and
41 relates them to climate change, is a major shortcoming, as it ignores other cases where conflicts did
42 not occur (Adams et al., 2018) despite degradation of the natural resource base and extreme weather
43 events.

45 **3.4.2.9 Impacts on Migration**

46 Environmentally-induced migration is complex and accounts for multiple drivers of mobility as well
47 as other adaptation measures undertaken by populations exposed to environmental risk (*high*
48 *confidence*). There is *medium evidence* and *low agreement* that climate change impacts migration. The
49 World Bank (2018) predicted that 143 million people would be forced to move internally by 2050 if

1 no climate action is taken. Focusing on asylum seekers alone, rather than the total number of
2 migrants, Missirian and Schlenker (2017) predict the asylum applications to the European Union will
3 increase from 28% (98,000 additional asylum applications per year) up to 188% (660,000 additional
4 applications per year) depending on the climate scenario by 2100. While the modelling efforts have
5 greatly improved over the years (Hunter et al., 2015; McLeman, 2011; Sherbinin and Bai, 2018) and
6 in particular, these recent estimates provide an important insight into potential future developments,
7 the quantitative projections are still based on the number of people exposed to risk rather than the
8 number of people who would actually engage in migration as a response to this risk (Gemenne, 2011;
9 McLeman, 2013) and they do not take into account individual agency in migration decision nor
10 adaptive capacities of individuals (Hartmann, 2010; Kniveton et al., 2011; Piguet, 2010) (see Section
11 3.6.2 discussing migration as a response to desertification). Accordingly, the available micro-level
12 evidence suggests that climate-related shocks are one of the many drivers of migration (Adger et al.,
13 2014; London Government Office for Science and Foresight, 2011; Melde et al., 2017), but the
14 individual responses to climate risk are more complex than commonly assumed (Gray and Mueller,
15 2012a). For example, despite strong focus on natural disasters, neither flooding (Gray and Mueller,
16 2012b; Mueller et al., 2014) nor earthquakes (Halliday, 2006) were found to induce long-term
17 migration; but instead, slow-onset changes, especially those provoking crop failures and heat stress,
18 could affect household or individual migration decisions (Gray and Mueller, 2012a; Missirian and
19 Schlenker, 2017; Mueller et al., 2014). Out-migration from drought-prone areas has received
20 particular attention (de Sherbinin et al., 2012; Ezra and Kiros, 2001) and indeed, a substantial body of
21 literature suggests that households engage in local or internal migration as a response to drought
22 (Findlay, 2011; Gray and Mueller, 2012a), while international migration decreases with drought in
23 some contexts (Henry et al., 2004), but might increase in contexts where migration networks are well
24 established (Feng et al., 2010; Nawrotzki and DeWaard, 2016; Nawrotzki et al., 2015, 2016).
25 Similarly, the evidence is not conclusive with respect to the effect of environmental drivers, in
26 particular desertification, on mobility. While it has not consistently entailed out-migration in the case
27 of Ecuadorian Andes (Gray, 2009, 2010) environmental and land degradation increased mobility in
28 Kenya and Nepal (Gray, 2011; Massey et al., 2010), but marginally decreased mobility in Uganda
29 (Gray, 2011). These results suggest that in some contexts, environmental shocks actually undermine
30 household's financial capacity to undertake migration (Nawrotzki and Bakhtsiyarava, 2017),
31 especially in the case of the poorest households (Barbier and Hochard, 2018; Koubi et al., 2016;
32 Kubik and Maurel, 2016; McKenzie and Yang, 2015). Adding to the complexity, migration,
33 especially to frontier areas, by increasing pressure on land and natural resources, might itself
34 contribute to environmental degradation at the destination (Hugo, 2008; IPBES, 2018a; McLeman,
35 2017). The consequences of migration can also be salient in the case of migration to urban or peri-
36 urban areas; indeed, environmentally-induced migration can add to urbanisation (3.6.2.2), often
37 exacerbating problems related to poor infrastructure and unemployment.

38 ***3.4.2.10 Impacts on Pastoral Communities***

39 Pastoral production systems occupy a significant portion of the world (Rass, 2006; Dong, 2016). Food
40 insecurity among pastoral households is often high (3.1.3; Gomes, 2006). The Sahelian droughts of
41 the 1970-80s provided an example of how droughts could affect livestock resources and crop
42 productivity, contributing to hunger, out-migration and suffering for millions of pastoralists (Hein and
43 De Ridder, 2006; Molua and Lambi, 2007). During these Sahelian droughts low and erratic rainfall
44 exacerbated desertification processes, leading to ecological changes that forced people to use marginal
45 lands and ecosystems. Similarly, the rate of rangeland degradation is now increasing because of
46 environmental changes and overexploitation of resources (Kassahun et al., 2008; Vetter, 2005).
47 Desertification coupled with climate change is negatively affecting livestock feed and grazing species
48 (Hopkins and Del Prado, 2007), changing the composition in favour of species with low forage
49 quality, ultimately reducing livestock productivity (D'Odorico et al., 2013; Dibari et al., 2016) and

1 increasing livestock disease prevalence (Thornton et al., 2009). There is *robust evidence and high*
2 *agreement* that weak adaptive capacity, coupled with negative effects from other climate-related
3 factors, are predisposing pastoralists to increased poverty from desertification and climate change
4 globally (López-i-Gelats et al., 2016; Giannini et al., 2008; IPCC, 2007). On the other hand,
5 misguided policies such as enforced sedentarisation and in certain cases protected area delineation
6 (fencing), which restrict livestock mobility have hampered optimal use of grazing land resources (Du,
7 2012); and led to degradation of resources and out-migration of people in search of better livelihoods
8 (Gebeye, 2016; Liao et al., 2015). Restrictions on the mobile lifestyle is reducing the resilient adaptive
9 capacity of pastoralists to natural hazards including extreme and variable weather conditions, drought
10 and climate change (Schilling et al., 2014). Furthermore, the exacerbation of the desertification
11 phenomenon due to agricultural intensification (D’Odorico et al., 2013) and land fragmentation
12 caused by encroachment of agriculture into rangelands (Otuoma et al., 2009; Behnke and Kerven,
13 2013) is threatening pastoral livelihoods. For example, commercial cotton (*Gossypium hirsutum*)
14 production is crowding out pastoral systems in Benin (Tamou et al., 2018). Food shortages and the
15 urgency to produce enough crop for public consumption are leading to the encroachment of
16 agriculture into productive rangelands and those converted rangelands are frequently prime lands used
17 by pastoralists to produce feed and graze their livestock during dry years (Dodd, 1994). The
18 sustainability of pastoral systems is therefore coming into question because of social and political
19 marginalisation of the system (Davies et al., 2016) and also because of the fierce competition it is
20 facing from other livelihood sources such as crop farming (Haan et al., 2016).

21 22 **3.5. Future Projections**

23 **3.5.1. Future Projections of Desertification**

24 Assessing the impact of climate change on future desertification is difficult as several environmental
25 and anthropogenic variables interact to determine its dynamics. The majority of modelling studies
26 regarding the future evolution of desertification rely on the analysis of specific climate change
27 scenarios and Global Climate Models and their effect on a few processes or drivers that trigger
28 desertification (Cross-Chapter Box 1: Scenarios, Chapter 1).

29 With regards to climate impacts, the analysis of global and regional climate models concludes that
30 under all representative concentration pathways (RCPs) potential evapotranspiration (PET) would
31 increase worldwide as a consequence of increasing surface temperatures and surface water vapour
32 deficit (Sherwood and Fu, 2014). Consequently, there would be associated changes in aridity indices
33 that depend on this variable (*high agreement, robust evidence*) (Cook et al., 2014a; Dai, 2011;
34 Dominguez et al., 2010; Feng and Fu, 2013; Ficklin et al., 2016; Fu et al., 2016; Greve and
35 Seneviratne, 1999; Koutroulis, 2019; Scheff and Frierson, 2015). Due to the large increase in PET and
36 decrease in precipitation over some subtropical land areas, aridity index will decrease in some
37 drylands (Zhao and Dai, 2015), with one model estimating ~10% increase in hyper-arid areas globally
38 (Zeng and Yoon, 2009). Increases in PET are projected to continue due to climate change (Cook et al.,
39 2014a; Fu et al., 2016; Lin et al., 2015; Scheff and Frierson, 2015). However, as noted in sections
40 3.1.1 and 3.2.1, these PET calculations use assumptions that are not valid in an environment with
41 changing CO₂. Evidence from precipitation, runoff or photosynthetic uptake of CO₂ suggest that a
42 future warmer world will be less arid (Roderick et al., 2015). Observations in recent decades indicate
43 that the Hadley cell has expanded poleward in both hemispheres (Fu et al., 2006; Hu and Fu, 2007;
44 Johanson et al., 2009; Seidel and Randel, 2007), and under all RCPs would continue expanding
45 (Johanson et al., 2009; Lu et al., 2007). This expansion leads to the poleward extension of sub-tropical
46 dry zones and hence an expansion in drylands on the poleward edge (Scheff and Frierson, 2012).

1 Overall, this suggests that while aridity will increase in some places (*high confidence*), there is
2 insufficient evidence to suggest a global change in dryland aridity (*medium confidence*).

3 Regional modelling studies confirm the outcomes of Global Climate Models (Africa: Terink et al.,
4 2013; China: Yin et al., 2015; Brazil: Marengo and Bernasconi, 2015; Cook et al., 2012; Greece:
5 Nastos et al., 2013; Italy: Coppola and Giorgi, 2009). According to the IPCC AR5 (IPCC, 2013),
6 decreases in soil moisture are detected in the Mediterranean, Southwest USA and southern African
7 regions. This is in line with alterations in the Hadley circulation and higher surface temperatures. This
8 surface drying will continue to the end of this century under the RCP8.5 scenario (*high confidence*).
9 Ramarao et al., (2015) showed that a future climate projection based on RCP4.5 scenario indicated the
10 possibility for detecting the summer-time soil drying signal over the Indian region during the 21st
11 century in response to climate change. The IPCC Special Report on Global Warming of 1.5°C (SR15,
12 Chapter 3) (Hoegh-Guldberg et al., 2018) report concluded with “*medium confidence*” that global
13 warming by more than 1.5°C increases considerably the risk of aridity for the Mediterranean area and
14 Southern Africa. Miao et al., (2015b) showed an acceleration of desertification trends under the
15 RCP8.5 scenario in the middle and northern part of Central Asia and some parts of north western
16 China. It is also useful to consider the effects of the dynamic–thermodynamical feedback of the
17 climate. Schewe and Levermann (2017) show increases up to 300 % in the central Sahel rainfall by
18 the end of the century due to an expansion of the West African monsoon. Warming could trigger an
19 intensification of monsoonal precipitation due to increases in ocean moisture availability.

20 The impacts of climate change on dust storm activity are not yet comprehensively studied and
21 represent an important knowledge gap. Currently, GCMs are unable to capture recent observed dust
22 emission and transport (Evan, 2018; Evan et al., 2014) limiting confidence in future projections.
23 Literature suggests that climate change decreases wind erosion/dust emission overall with regional
24 variation (*low confidence*). Mahowald et al. (2006) and Mahowald (2007) found that climate change
25 led to a decrease in desert dust source areas globally using CMIP3 GCMs. Wang et al. (2009) found a
26 decrease in sand dune movement by 2039 (increasing thereafter) when assessing future wind erosion
27 driven desertification in arid and semiarid China using a range of SRES scenarios and HadCM3
28 simulations. Dust activity in the US Southern Great Plains was projected to increase, while in the
29 Northern Great Plains it was projected to decrease under RCP 8.5 climate change scenario (Pu and
30 Ginoux, 2017). Evan et al. (2016) project a decrease in African dust emission associated with a
31 slowdown of the tropical circulation in the high CO₂ RCP8.5 scenario.

32 Global estimates of the impact of climate change on soil salinisation show that under the IS92a
33 emissions scenario (a scenario prepared in 1992 that contains “business as usual” assumptions)
34 (Leggett et al., 1992) the area at risk of salinisation would increase in the future (*limited evidence*,
35 *high agreement*; Schofield and Kirkby, 2003). Climate change has an influence on soil salinisation
36 that induces further land degradation through several mechanisms that vary in their level of
37 complexity. However, only a few examples can be found to illustrate this range of impacts, including
38 the effect of groundwater table depletion (Rengasamy, 2006) and irrigation management (Sivakumar,
39 2007), salt migration in coastal aquifers with decreasing water tables (Sherif and Singh, 1999; 4.10.7),
40 and surface hydrology and vegetation that affect wetlands and favour salinisation (Nielsen and Brock,
41 2009).

42

43 **3.5.1.1. Future Vulnerability and Risk to Desertification**

44 Following the conceptual framework developed in the Managing the Risks of Extreme Events and
45 Disasters to Advance Climate Change Adaptation special report (SREX) (IPCC, 2012), future risks
46 are assessed by examining changes in exposure (i.e. presence of people; livelihoods; species or
47 ecosystems; environmental functions, service, and resources; infrastructure; or economic, social or
48 cultural assets; see glossary), changes in vulnerability (i.e. propensity or predisposition to be

1 adversely affected; see glossary) and changes in the nature and magnitude of hazards (i.e. potential
2 occurrence of a natural or human-induced physical event that causes damage; see glossary). Climate
3 change is expected to further exacerbate the vulnerability of dryland ecosystems to desertification by
4 increasing PET globally (Sherwood and Fu, 2014). Temperature increases between 2°C and 4°C are
5 projected in drylands by the end of the 21st century under RCP4.5 and RCP8.5 scenarios, respectively
6 (IPCC, 2013). An assessment by (Carrão et al., 2017) showed an increase in drought hazards by late-
7 century (2071–2099) compared to a baseline (1971–2000) under high RCPs in drylands around the
8 Mediterranean, south-eastern Africa, and southern Australia. In Latin America, Morales et al. (2011)
9 indicated that areas affected by drought will increase significantly by 2100 under SRES scenarios A2
10 and B2. The countries expected to be affected include Guatemala, El Salvador, Honduras and
11 Nicaragua. In CMIP5 scenarios, Mediterranean types of climate are projected to become drier
12 (Alessandri et al., 2014; Polade et al., 2017), with the equatorward margins being potentially replaced
13 by arid climate types (Alessandri et al., 2014). Globally, climate change is predicted to intensify the
14 occurrence and severity of droughts (*medium confidence*) (2.2; Dai, 2013; Sheffield and Wood, 2008;
15 Swann et al., 2016; Wang, 2005; Zhao and Dai, 2015; Carrão et al., 2017; Naumann et al., 2018).
16 Ukkola et al. (2018) showed large discrepancies between CMIP5 models for all types of droughts,
17 limiting the confidence that can be assigned to projections of drought.

18 Drylands are characterised by high climatic variability. Climate impacts on desertification are not
19 only defined by projected trends in mean temperature and precipitation values but are also strongly
20 dependent on changes in climate variability and extremes (Reyer et al., 2013). The responses of
21 ecosystems depend on diverse vegetation types. Drier ecosystems are more sensitive to changes in
22 precipitation and temperature (Li et al., 2018; Seddon et al., 2016; You et al., 2018), increasing
23 vulnerability to desertification. It has also been reported that areas with high variability in
24 precipitation tend to have lower livestock densities and that those societies that have a strong
25 dependence on livestock that graze natural forage are especially affected (Sloat et al., 2018). Social
26 vulnerability in drylands increases as a consequence of climate change that threatens the viability of
27 pastoral food systems (Dougill et al., 2010; López-i-Gelats et al., 2016). Social drivers can also play
28 an important role with regards to future vulnerability (Máñez Costa et al., 2011). In the arid region of
29 north-western China, Liu et al. (2016b) estimated that under RCP4.5 areas of increased vulnerability
30 to climate change and desertification will surpass those with decreased vulnerability.

31 Using an ensemble of global climate, integrated assessment and impact models, Byers et al. (2018)
32 investigated 14 impact indicators at different levels of global mean temperature change and
33 socioeconomic development. The indicators cover water, energy and land sectors. Of particular
34 relevance to desertification are the water (e.g. water stress, drought intensity) and the land (e.g. habitat
35 degradation) indicators. Under shared socioeconomic pathway SSP2 (“Middle of the Road”) at 1.5°C,
36 2°C and 3°C of global warming, the numbers of dryland populations exposed (vulnerable) to various
37 impacts related to water, energy and land sectors (e.g. water stress, drought intensity, habitat
38 degradation) are projected to reach 951 (178) million, 1,152 (220) million and 1,285 (277) million,
39 respectively. While at global warming of 2°C, under SSP1 (sustainability), the exposed (vulnerable)
40 dryland population is 974 (35) million, and under SSP3 (Fragmented World) it is 1,267 (522) million.
41 Steady increases in the exposed and vulnerable populations are seen for increasing global mean
42 temperatures. However much larger differences are seen in the vulnerable population under different
43 SSPs. Around half the vulnerable population is in South Asia, followed by Central Asia, West Africa
44 and East Asia.

45 **3.5.2. Future Projections of Impacts**

46 Future climate change is expected to increase the potential for increased soil erosion by water in
47 dryland areas (*medium confidence*). Yang et al. (2003) use a Revised Universal Soil Loss Equation
48 (RUSLE) model to study global soil erosion under historical, present and future conditions of both

1 cropland and climate. Soil erosion potential has increased by about 17%, and climate change will
2 increase this further in the future. In northern Iran, under the SRES A2 emission scenario the mean
3 erosion potential is projected to grow by 45% comparing the period 1991-2010 with 2031-2050 (Zare
4 et al., 2016). A strong decrease in precipitation for almost all parts of Turkey was projected for the
5 period of 2021–2050 compared to 1971-2000 using Regional Climate Model, RegCM4.4 of the
6 International Centre for Theoretical Physics (ICTP) under RCP4.5 and RCP8.5 scenarios (Türkeş et
7 al., 2019). The projected changes in precipitation distribution can lead to more extreme precipitation
8 events and prolonged droughts, increasing Turkey’s vulnerability to soil erosion (Türkeş et al.,
9 2019). In Portugal, a study comparing wet and dry catchments under A1B and B1 emission scenarios
10 showed an increase in erosion in dry catchments (Serpa et al., 2015). In Morocco an increase in
11 sediment load is projected as a consequence of reduced precipitation (Simonneaux et al., 2015). WGII
12 AR5 concluded the impact of increases in heavy rainfall and temperature on soil erosion will be
13 modulated by soil management practices, rainfall seasonality and land cover (Jiménez Cisneros et al.,
14 2014). Ravi et al. (2010) predicted an increase in hydrologic and aeolian soil erosion processes as a
15 consequence of droughts in drylands. However, there are some studies that indicate that soil erosion
16 will be reduced in Spain (Zabaleta et al., 2013), Greece (Nerantzaki et al., 2015) and Australia (Klik
17 and Eitzinger, 2010), while others project changes in erosion as a consequence of the expansion of
18 croplands (Borrelli et al., 2017).

19 Potential dryland expansion implies lower C sequestration and higher risk of desertification (Huang et
20 al., 2017), with severe impacts on land usability and threatening food security. At the level of biomes
21 (global-scale zones, generally defined by the type of plant life that they support in response to average
22 rainfall and temperature patterns; see glossary), soil C uptake is determined mostly by weather
23 variability. The area of the land in which dryness controls CO₂ exchange has risen by 6% since 1948
24 and is projected to expand by at least another 8% by 2050. In these regions net C uptake is about 27%
25 lower than elsewhere (Yi et al., 2014). Potential losses of soil C are projected to range from 9 to 12%
26 of the total C stock in the 0-20 cm layer of soils in the southern European Russia by end of this
27 century (Ivanov et al., 2018).

28 Desertification under climate change will threaten biodiversity in drylands (*medium confidence*).
29 Rodriguez-Rodriguez-Caballero et al. (2018) analysed the cover of biological soil crusts under current
30 and future environmental conditions utilising an environmental niche modelling approach. Their
31 results suggest that biological soil crusts currently cover ~1600 M ha in drylands. Under RCP
32 scenarios 2.6 to 8.5, 25–40% of this cover will be lost by 2070 with climate and land use contributing
33 equally. The predicted loss is expected to substantially reduce their contribution to N cycling (6.7–
34 9.9 Tg yr⁻¹ of N) and C cycling (0.16–0.24 Pg yr⁻¹ of C) (Rodriguez-Caballero et al., 2018). A study
35 in Colorado Plateau, USA showed that changes in climate in drylands may damage the biocrust
36 communities by promoting rapid mortality of foundational species (Rutherford et al., 2017), while in
37 southern California deserts climate change-driven extreme heat and drought may surpass the survival
38 thresholds of some desert species (Bachelet et al., 2016). In semiarid Mediterranean shrublands in
39 eastern Spain, plant species richness and plant cover could be reduced by climate change and soil
40 erosion (García-Fayos and Bochet, 2009). The main drivers of species extinctions are land use
41 change, habitat pollution, over-exploitation, and species invasion, while the climate change is
42 indirectly linked to species extinctions (Settele et al., 2014). Malcolm et al. (2006) found that more
43 than 2000 plant species located within dryland biodiversity hotspots could become extinct within 100
44 years starting 2004 (within the Cape Floristic Region, Mediterranean Basin and Southwest Australia).
45 Furthermore, it is suggested that land use and climate change could cause the loss of 17% of species
46 within shrublands and 8% within hot deserts by 2050 (van Vuuren et al., 2006) (*low confidence*). A
47 study in the semi-arid Chinese Altai Mountains showed that mammal species richness will decline,
48 and rates of species turnover will increase, and more than 50% of their current ranges will be lost (Ye
49 et al., 2018).

1 Changing climate and land use have resulted in higher aridity and more droughts in some drylands,
2 with the rising role of precipitation, wind and evaporation on desertification (Fischlin et al., 2007). In
3 a 2°C world, annual water discharge is projected to decline, and heatwaves are projected to pose risk
4 to food production by 2070 (Waha et al., 2017). However, Betts et al. (2018) found a mixed response
5 of water availability (runoff) in dryland catchments to global temperature increases from 1.5°C to 2°C.
6 The forecasts for Sub-Saharan Africa suggest that higher temperatures, increase in the number of
7 heatwaves, and increasing aridity, will affect the rainfed agricultural systems (Serdeczny et al., 2017).
8 A study by (Wang et al., 2009) in arid and semiarid China showed decreased livestock productivity
9 and grain yields from 2040 to 2099, threatening food security. In Central Asia, projections indicate a
10 decrease in crop yields, and negative impacts of prolonged heat waves on population health (3.7.2;
11 Reyner et al., 2017). World Bank (2009) projected that, without the C fertilisation effect, climate
12 change will reduce the mean yields for 11 major global crops, such as millet, field pea, sugar beet,
13 sweet potato, wheat, rice, maize, soybean, groundnut, sunflower, and rapeseed, by 15% in Sub-
14 Saharan Africa, 11% in Middle East and North Africa, 18% in South Asia, and 6% in Latin America
15 and Caribbean by 2046–2055, compared to 1996–2005. A separate meta-analysis suggested a similar
16 reduction in yields in Africa and South Asia due to climate change by 2050 (Knox et al., 2012).
17 Schlenker and Lobell (2010) estimated that in sub-Saharan Africa, crop production may be reduced by
18 17–22% due to climate change by 2050. At the local level, climate change impacts on crop yields vary
19 by location (5.2.2). Negative impacts of climate change on agricultural productivity contribute to
20 higher food prices. The imbalance between supply and demand for agricultural products is projected
21 to increase agricultural prices in the range of 31% for rice to 100% for maize by 2050 (Nelson et al.,
22 2010), and cereal prices in the range between a 32% increase and a 16% decrease by 2030 (Hertel
23 et al., 2010). In the southern European Russia, it is projected that the yields of grain crops will decline
24 by 5 to 10% by 2050 due to the higher intensity and coverage of droughts (Ivanov et al., 2018).

25
26 Climate change can have strong impacts on poverty in drylands (*medium confidence*) (Hallegatte and
27 Rozenberg, 2017; Hertel and Lobell, 2014). Globally, Hallegatte et al. (2015) project that without
28 rapid and inclusive progress on eradicating multidimensional poverty, climate change can increase the
29 number of the people living in poverty by 35 to 122 million people until 2030. Although these
30 numbers are global and not specific to drylands, the highest impacts in terms of the share of the
31 national populations being affected are projected to be in the drylands areas of the Sahel region,
32 eastern Africa and South Asia (Stephane Hallegatte et al., 2015). The impacts of climate change on
33 poverty vary depending on whether the household is a net agricultural buyer or seller. Modelling
34 results showed that poverty rates would increase by about one-third among the urban households and
35 non-agricultural self-employed in Malawi, Uganda, Zambia, and Bangladesh due to high agricultural
36 prices and low agricultural productivity under climate change (Hertel et al., 2010). On the contrary,
37 modelled poverty rates fell substantially among agricultural households in Chile, Indonesia,
38 Philippines and Thailand, because higher prices compensated for productivity losses (Hertel et al.,
39 2010).

40 41 **3.6. Responses to Desertification under Climate Change**

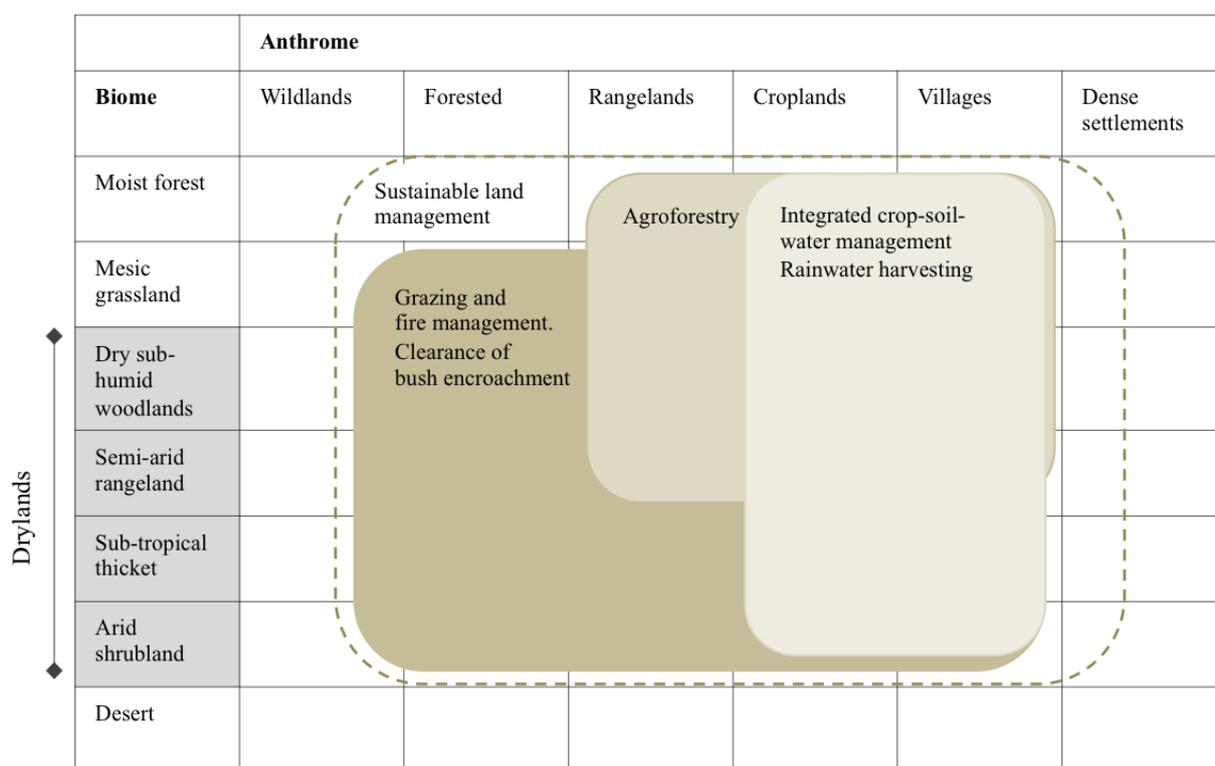
42 Achieving sustainable development of dryland livelihoods requires avoiding dryland degradation
43 through SLM and restoring and rehabilitating the degraded drylands due to their potential wealth of
44 ecosystem benefits and importance to human livelihoods and economies (Thomas, 2008). A broad
45 suite of on the ground response measures exist to address desertification (Scholes, 2009), be it in the
46 form of improved fire and grazing management, the control of erosion; integrated crop, soil and water
47 management, among others (Liniger and Critchley, 2007; Scholes, 2009). These actions are part of the
48 broader context of dryland development and long-term SLM within coupled socio-economic systems
49 (Reynolds et al., 2007; Stringer et al., 2017; Webb et al., 2017). Many of these response options

1 correspond to those grouped under land transitions in the IPCC Special Report on Global Warming of
 2 1.5°C (Table 6.4; Coninck et al., 2018). It is therefore recognised that such actions require financial,
 3 institutional and policy support for their wide-scale adoption and sustainability over time (3.6.3; 4.8.5;
 4 6.4.4).

5 **3.6.1. SLM Technologies and Practices: on the Ground Actions**

6 A broad range of activities and measures can help avoid, reduce and reverse degradation across the
 7 dryland areas of the world. Many of these actions also contribute to climate change adaptation and
 8 mitigation, with further sustainable development co-benefits for poverty reduction and food security
 9 (*high confidence*) (6.3). As preventing desertification is strongly preferable and more cost-effective
 10 than allowing land to degrade and then attempting to restore it (IPBES, 2018b; Webb et al., 2013),
 11 there is a growing emphasis on avoiding and reducing land degradation, following the Land
 12 Degradation Neutrality framework (Cowie et al., 2018; Orr et al., 2017; 4.8.5).

13



14

15 **Figure 3.10 The typical distribution of on-the-ground actions across global biomes and anthromes**

16

17 An assessment is made of six activities and measures practicable across the biomes and anthromes of
 18 the dryland domain (Figure 3.10). This suite of actions is not exhaustive, but rather a set of activities
 19 that are particularly pertinent to global dryland ecosystems. They are not necessarily exclusive to
 20 drylands and are often implemented across a range of biomes and anthromes (Figure 3.10). For
 21 afforestation, see 3.7.2, Cross-Chapter Box 2 in Chapter 1 and Chapter 4 (4.8.3). The use of
 22 anthromes as a structuring element for response options is based on the essential role of interactions
 23 between social and ecological systems in driving desertification within coupled socio-ecological
 24 systems (Cherlet et al., 2018). The concept of the anthromes is defined in the glossary and explored
 25 further in Chapters 1, 4 and 6.

1 The assessment of each action is twofold: firstly, to assess the ability of each action to address
2 desertification and enhance climate change resilience, and secondly, to assess the potential impact of
3 future climate change on the effectiveness of each action.

4 **3.6.1.1. Integrated Crop-Soil-Water Management**

5 Forms of integrated cropland management have been practiced in drylands for over thousands of
6 years (Knörzer et al., 2009). Actions include planting a diversity of species including drought tolerant
7 crops, reducing tillage, applying organic compost and fertiliser, adopting different forms of irrigation
8 and maintaining vegetation and mulch cover. In the contemporary era, several of these actions have
9 been adopted in response to climate change.

10 In terms of climate change *adaptation*, the resilience of agriculture to the impacts of climate change is
11 strongly influenced by the underlying health and stability of soils as well as improvements in crop
12 varieties, irrigation efficiency and supplemental irrigation, e.g. through rainwater harvesting (*medium*
13 *evidence, high agreement*, Altieri et al., 2015; Amundson et al., 2015; Derpsch et al., 2010; Lal, 1997;
14 de Vries et al., 2012). Desertification often leads to a reduction in ground cover that in turn results in
15 accelerated water and wind erosion and an associated loss of fertile topsoil that can greatly reduce the
16 resilience of agriculture to climate change (*medium evidence, high agreement*, (Touré et al., 2019;
17 Amundson et al., 2015; Borrelli et al., 2017; Pierre et al., 2017). Amadou et al. (2011) note that even a
18 minimal cover of crop residues (100 kg ha⁻¹) can substantially decrease wind erosion.

19 Compared to conventional (flood or furrow) irrigation, drip irrigation methods are more efficient in
20 supplying water to the plant root zone, resulting in lower water requirements and enhanced water use
21 efficiency (*robust evidence and high agreement*) (Ibragimov et al., 2007; Narayanamoorthy, 2010;
22 Niaz et al., 2009). For example, in the rainfed area of Fetejjang, Pakistan, the adoption of drip
23 methods reduced water usage by 67-68% during the production of tomato, cucumber and bell peppers,
24 resulting in a 68-79% improvement in water use efficiency compared to previous furrow irrigation
25 (Niaz et al., 2009). In India, drip irrigation reduced the amount of water consumed in the production
26 of sugarcane by 44%, grapes by 37%, bananas by 29% and cotton by 45%, while enhancing the yields
27 by up to 29% (Narayanamoorthy, 2010). Similarly, in Uzbekistan, drip irrigation increased the yield
28 of cotton by 10-19% while reducing water requirements by 18-42% (Ibragimov et al., 2007).

29 A prominent response that addresses soil loss, health and cover is altering cropping methods. The
30 adoption of intercropping (inter- and intra- row planting of companion crops) and relay cropping
31 (temporally differentiated planting of companion crops) maintains soil cover over a larger fraction of
32 the year, leading to an increase in production, soil N, species diversity and a decrease in pest
33 abundance (*robust evidence and medium agreement*, (Altieri and Koochafkan, 2008; Tanveer et al.,
34 2017; Wilhelm and Wortmann, 2004). For example, intercropping maize and sorghum with
35 *Desmodium* (an insect repellent forage legume) and *Brachiaria* (an insect trapping grass), which is
36 being promoted in drylands of East Africa, led to a two-three fold increase in maize production and
37 an 80% decrease in stem boring insects (Khan et al., 2014). In addition to changes in cropping
38 methods, forms of agroforestry and shelter belts are often used to reduce erosion and improve soil
39 conditions (3.7.2). For example, the use of tree belts of mixed species in northern China led to a
40 reduction of surface wind speed and an associated reduction in soil temperature by up to 40% and an
41 increase in soil moisture by up to 30% (Wang et al., 2008).

42 A further measure that can be of increasing importance under climate change is rainwater harvesting
43 (RWH), including traditional zai (small basins used to capture surface runoff), earthen bunds and
44 ridges (Nyamadzawo et al., 2013), *fanya juu* infiltration pits (Nyagumbo et al., 2019), contour stone
45 bunds (Garrity et al., 2010) and semi-permeable stone bunds (often referred to by the French term
46 "digue filtrante") (Taye et al., 2015). RWH increases the amount of water available for agriculture and
47 livelihoods through the capture and storage of runoff, while at the same time, reducing the intensity of
48 peak flows following high intensity rainfall events. It is therefore often highlighted as a practical

1 response to dryness (i.e. long-term aridity and low seasonal precipitation) and rainfall variability
2 projected to become more acute over time in some dryland areas (Dile et al., 2013; Vohland and
3 Barry, 2009). For example, for Wadi Al-Lith drainage in Saudi Arabia, the use of rainwater harvesting
4 was suggested as a key climate change adaptation action (Almazroui et al., 2017). There is *robust*
5 *evidence and high agreement* that the implementation of RWH systems leads to an increase in
6 agricultural production in drylands (see reviews by Biazin et al., 2012; Bouma and Wösten, 2016;
7 Dile et al., 2013). A global meta-analysis of changes in crop production due to the adoption of RWH
8 techniques noted an average increase in yields of 78%, ranging from –28% to 468% (Bouma and
9 Wösten, 2016). Of particular relevance to climate change in drylands is that the relative impact of
10 RWH on agricultural production generally increases with increasing dryness. Relative yield
11 improvements due to the adoption of RWH were significantly higher in years with less than 330 mm
12 rainfall, compared to years with more than 330 mm (Bouma and Wösten, 2016). Despite delivering a
13 clear set of benefits, there are some issues that need to be considered. The impact RWH may vary at
14 different temporal and spatial scales (Vohland and Barry, 2009). At a plot scale, RWH structures may
15 increase available water and enhance agricultural production, SOC and nutrient availability, yet at a
16 catchment scale, they may reduce runoff to downstream uses (Meijer et al., 2013; Singh et al., 2012;
17 Vohland and Barry, 2009; Yosef and Asmamaw, 2015). Inappropriate storage of water in warm
18 climates can lead to an increase in water related diseases unless managed correctly, for example,
19 schistosomiasis and malaria (see review by Boelee et al., 2013).

20 Integrated crop-soil-water management may also deliver climate change *mitigation* benefits through
21 avoiding, reducing and reversing the loss of SOC (Table 6.5). Approximately 20-30 Pg of SOC have
22 been released into the atmosphere through desertification processes, for example, deforestation,
23 overgrazing and conventional tillage (Lal, 2004). Activities, such as those associated with
24 conservation agriculture (minimising tillage, crop rotation, maintaining organic cover and planting a
25 diversity of species), reduce erosion, improve water use efficiency and primary production, increase
26 inflow of organic material and enhance SOC over time, contributing to climate change mitigation
27 and adaptation (*high confidence*) (Plaza-Bonilla et al., 2015; Lal, 2015; Srinivasa Rao et al., 2015;
28 Sombrero and de Benito, 2010). Conservation agriculture practices also lead to increases in SOC
29 (*medium confidence*). However, sustained C sequestration is dependent on net primary productivity
30 and on the availability of crop-residues that may be relatively limited and often consumed by
31 livestock or used elsewhere in dryland contexts (Cheesman et al., 2016; Plaza-Bonilla et al., 2015).
32 For this reason, expected rates of C sequestration following changes in agricultural practices in
33 drylands are relatively low (0.04-0.4 t C ha⁻¹) and it may take a protracted period of time, even several
34 decades, for C stocks to recover if lost (*medium confidence*) (Farage et al., 2007; Hoyle, D'Antuono,
35 Overheu, and Murphy, 2013; Lal, 2004). This long recovery period enforces the rationale for
36 prioritising avoiding and reducing land degradation and loss of C, in addition to restoration activities.

37

38 **3.6.1.2. Grazing and Fire Management in Drylands**

39 Rangeland management systems such as sustainable grazing approaches and re-vegetation increase
40 rangeland productivity (*high confidence*) (Table 6.5). Open grassland, savanna and woodland are
41 home to the majority of world's livestock production (Safriel et al., 2005). Within these drylands
42 areas, prevailing grazing and fire regimes play an important role in shaping the relative abundance of
43 trees versus grasses (Scholes and Archer, 1997; Staver et al., 2011; Stevens et al., 2017), as well as
44 the health of the grass layer in terms of primary production, species richness and basal cover (the
45 proportion of the plant that is in the soil) (Plaza-Bonilla et al., 2015; Short et al., 2003). This in turn
46 influences levels of soil erosion, soil nutrients, secondary production and additional ecosystem
47 services (Divinsky et al., 2017; Pellegrini et al., 2017). A further set of drivers, including soil type,
48 annual rainfall and changes in atmospheric CO₂ may also define observed rangeland structure and

1 composition (Devine et al., 2017; Donohue et al., 2013), but the two principal factors that pastoralists
2 can manage are grazing and fire by altering their frequency, type and intensity.

3
4 The impact of grazing and fire regimes on biodiversity, soil nutrients, primary production and further
5 ecosystem services is not constant and varies between locations (Divinsky et al., 2017; Fleischner,
6 1994; van Oijen et al., 2018). Trade-offs may therefore need to be considered to ensure that rangeland
7 diversity and production are resilient to climate change (Plaza-Bonilla et al., 2015; van Oijen et al.,
8 2018). In certain locations, even light to moderate grazing have led to a significant decrease in the
9 occurrence of particular species, especially forbs (O'Connor et al., 2011; Scott-shaw and Morris,
10 2015). In other locations, species richness is only significantly impacted by heavy grazing and is able
11 to withstand light to moderate grazing (Divinsky et al., 2017). A context specific evaluation of how
12 grazing and fire impact particular species may therefore be required to ensure the persistence of target
13 species over time (Marty, 2005). A similar trade-off may need to be considered between soil C
14 sequestration and livestock production. As noted by Plaza-Bonilla et al. (2015) increasing grazing
15 pressure has been found to both increase and decrease SOC stocks in different locations. Where it has
16 led to a decrease in soil C stocks, for example in Mongolia (Han et al., 2008) and Ethiopia (Bikila et
17 al., 2016), trade-offs between C sequestration and the value of livestock to local livelihoods need be
18 considered.

19
20 Although certain herbaceous species may be unable to tolerate grazing pressure, a complete lack of
21 grazing or fire may not be desired in terms of ecosystems health. It can lead to a decrease in basal
22 cover and the accumulation of moribund, unpalatable biomass that inhibits primary production
23 (Manson et al., 2007; Scholes, 2009). The utilisation of the grass sward through light to moderate
24 grazing stimulates the growth of biomass, basal cover and allows water services to be sustained over
25 time (Papanastasis et al., 2017; Scholes, 2009). Even, moderate to heavy grazing in periods of higher
26 rainfall may be sustainable, but constant heavy grazing during dry periods and especially droughts can
27 lead to a reduction in basal cover, SOC, biological soil crusts, ecosystem services and an accelerated
28 erosion (*high agreement, robust evidence*, (Archer et al., 2017; Conant and Paustian, 2003; D'Odorico
29 et al., 2013; Geist and Lambin, 2004; Havstad et al., 2006; Huang et al., 2007; Manzano and N avar,
30 2000; Pointing and Belnap, 2012; Weber et al., 2016). For this reason, the inclusion of drought
31 forecasts and contingency planning in grazing and fire management programs is crucial to avoid
32 desertification (Smith and Foran, 1992; Torell et al., 2010). It is an important component of avoiding
33 and reducing early degradation. Although grasslands systems may be relatively resilient and can often
34 recover from a moderately degraded state (Khishigbayar et al., 2015; Porensky et al., 2016), if a
35 tipping point has been exceeded, restoration to a historic state may not be economical or ecologically
36 feasible (D'Odorico et al., 2013).

37
38 Together with livestock management (Table 6.5), the use of fire is an integral part of rangeland
39 management and can be applied to remove moribund and unpalatable forage, exotic weeds and woody
40 species (Archer et al., 2017). Fire has less of an effect on SOC and soil nutrients in comparison to
41 grazing (Abril et al., 2005), yet elevated fire frequency has been observed to lead to a decrease in soil
42 C and N (Abril et al., 2005; Bikila et al., 2016; Bird et al., 2000; Pellegrini et al., 2017). Although the
43 impact of climate change on fire frequency and intensity may not be clear due to its differing impact
44 on fuel accumulation, suitable weather conditions and sources of ignition (Abatzoglou et al., 2018;
45 Littell et al., 2018; Moritz et al., 2012), there is an increasing use of prescribed fire to address several
46 global change phenomena, for example, the spread of invasive species and bush encroachment as well
47 as the threat of intense runaway fires (Fernandes et al., 2013; McCaw, 2013; van Wilgen et al., 2010).
48 Cross-Chapter Box 3 located in Chapter 2 provides a further review of the interaction between fire
49 and climate change.

1 There is often much emphasis on reducing and reversing the degradation of rangelands due to the
2 wealth of benefits they provide, especially in the context of assisting dryland communities to adapt to
3 climate change (Webb et al., 2017; Woollen et al., 2016). The emerging concept of ecosystem-based
4 adaptation has highlighted the broad range of important ecosystem services that healthy rangelands
5 can provide in a resilient manner to local residents and downstream economies (Kloos and Renaud,
6 2016; Reid et al., 2018). In terms of climate change mitigation, the contribution of rangelands,
7 woodland and sub-humid dry forest (e.g. Miombo woodland in south-central Africa) is often
8 undervalued due to relatively low C stocks per hectare. Yet due to their sheer extent, the amount of C
9 sequestered in these ecosystems is substantial and can make a valuable contribution to climate change
10 mitigation (Lal, 2004; Pelletier et al., 2018).

11 **3.6.1.3. Clearance of Bush Encroachment**

12 The encroachment of open grassland and savanna ecosystems by woody species has occurred for at
13 least the past 100 years (Archer et al., 2017; O'Connor et al., 2014; Schooley et al., 2018). Dependent
14 on the type and intensity of encroachment, it may lead to a net loss of ecosystem services and be
15 viewed as a form of desertification (Dougill et al., 2016; O'Connor et al., 2014). However, there are
16 circumstances where bush encroachment may lead to a net increase in ecosystem services, especially
17 at intermediate levels of encroachment, where the ability of the landscape to produce fodder for
18 livestock is retained, while the production of wood and associated products increases (Eldridge et al.,
19 2011; Eldridge and Soliveres, 2014). This may be particularly important in regions such as southern
20 Africa and India where over 65% of rural households depend on fuelwood from surrounding
21 landscapes as well as livestock production (Komala and Prasad, 2016; Makonese et al., 2017;
22 Shackleton and Shackleton, 2004).

23
24
25 This variable relationship between the level of encroachment, C stocks, biodiversity, provision of
26 water and pastoral value (Eldridge and Soliveres, 2014) can present a conundrum to policy makers,
27 especially when considering the goals of three Rio Conventions - UNFCCC, UNCCD and UNCBD.
28 Clearing intense bush encroachment may improve species diversity, rangeland productivity, the
29 provision of water and decrease desertification, thereby contributing to the goals of the UNCBD,
30 UNCCD as well as adaptation aims of the UNFCCC. However, it would lead to the release of biomass
31 C stocks into the atmosphere and potentially conflict with the mitigation aims of the UNFCCC.

32
33 For example, Smit et al. (2015) observed an average increase in above-ground woody C stocks of 44 t
34 C ha⁻¹ in savannas in northern Namibia. However, since bush encroachment significantly inhibited
35 livestock production, there are often substantial efforts to clear woody species (Stafford-Smith et al.,
36 2017). Namibia has an early national programme aimed at clearing woody species through
37 mechanical measures (harvesting of trees) as well as the application of arboricides (Smit et al., 2015).
38 However, the long-term success of clearance and subsequent improved fire and grazing management
39 remains to be evaluated, especially restoration back towards an 'original open grassland state'. For
40 example, in northern Namibia, the rapid reestablishment of woody seedlings has raised questions
41 about whether full clearance and restoration is possible (Smit et al., 2015). In arid landscapes, the
42 potential impact of elevated atmospheric CO₂ (Donohue et al., 2013; Kgope et al., 2010) and
43 opportunity to implement high intensity fires that remove woody species and maintain rangelands in
44 an open state has been questioned (Bond and Midgley, 2000). If these drivers of woody plant
45 encroachment cannot be addressed, a new form of "emerging ecosystem" (Milton, 2003) may need to
46 be explored that includes both improved livestock and fire management as well as the utilisation of
47 biomass as a long-term commodity and source of revenue (Smit et al., 2015). Initial studies in
48 Namibia and South Africa (Stafford-Smith et al., 2017) indicate that there may be good opportunity to
49 produce sawn timber, fencing poles, fuel wood and commercial energy, but factors such as the cost of
50 transport can substantially influence the financial feasibility of implementation.

1
2 The benefit of proactive management that prevents land from being degraded (altering grazing
3 systems or treating bush encroachment at early stages before degradation has been initiated) is more
4 cost-effective in the long-term and adds resistance to climate change than treating lands after
5 degradation has occurred (Webb et al., 2013; Weltz and Spaeth, 2012). The challenge is getting
6 producers to alter their management paradigm from short-term objectives to long-term objectives.
7

8 **3.6.1.4. Combating sand and dust storms through sand dune stabilisation**

9 Dust and sand storms have a considerable impact on natural and human systems (3.4.1, 3.4.2).
10 Application of sand dune stabilisation techniques contributes to reducing sand and dust storms (*high*
11 *confidence*). Using a number of methods, sand dune stabilisation aims to avoid and reduce the
12 occurrence of dust and sand storms (Mainguet and Dumay, 2011). Mechanical techniques include
13 building palisades to prevent the movement of sand and reduce sand deposits on infrastructure.
14 Chemical methods include the use of calcium bentonite or using silica gel to fix mobile sand
15 (Aboushook et al., 2012; Rammal and Jubair, 2015). Biological methods include the use of mulch to
16 stabilise surfaces (Sebaa et al., 2015; Yu et al., 2004) and establishing permanent plant cover using
17 pasture species that improve grazing at the same time (Abdelkebir and Ferchichi, 2015; Zhang et al.,
18 2015; 3.7.1.3). When the dune is stabilised, woody perennials are introduced that are selected
19 according to climatic and ecological conditions (FAO, 2011). For example, such revegetation
20 processes have been implemented on the shifting dunes of the Tengger Desert in northern China
21 leading to the stabilisation of sand and the sequestration of up to 10 t C ha⁻¹ over a period of 55 years
22 (Yang et al., 2014).

23 **3.6.1.5 Use of Halophytes for the Revegetation of Saline Lands**

24 Soil salinity and sodicity can severely limit the growth and productivity of crops (Jan et al., 2017) and
25 lead to a decrease in available arable land. Leaching and drainage provides a possible solution, but
26 can be prohibitively expensive. An alternative, more economical option, is the growth of halophytes
27 (plants that are adapted to grow under highly saline conditions) that allow saline land to be used in a
28 productive manner (Qadir et al., 2000). The biomass produced can be used as forage, food, feed,
29 essential oils, biofuel, timber, fuelwood (Chughtai et al., 2015; Mahmood et al., 2016; Sharma et al.,
30 2016). A further co-benefit is the opportunity to mitigate climate change through the enhancement of
31 terrestrial C stocks as land is revegetated (Dagar et al., 2014; Wicke et al., 2013). The combined use
32 of salt-tolerant crops, improved irrigation practices, chemical remediation measures and appropriate
33 mulch and compost is effective in reducing the impact of secondary salinisation (*medium confidence*).

34 In Pakistan, where about 6.2 M ha of agricultural land is affected by salinity, pioneering work on
35 utilising salt tolerant plants for the revegetation of saline lands (Biosaline Agriculture) was done in the
36 early 1970s (NIAB, 1997). A number of local and exotic varieties were initially screened for salt
37 tolerance in lab- and greenhouse based studies, and then distributed to similar saline areas (Ashraf et
38 al., 2010). These included tree species (*Acacia ampliceps*, *A. nilotica*, *Eucalyptus camaldulensis*,
39 *Prosopis juliflora*, *Azadirachta indica*) (Awan and Mahmood, 2017), forage plants (*Leptochloa fusca*,
40 *Sporobolus arabicus*, *Brachiaria mutica*, *Echinochloa* sp., *Sesbania* and *Atriplex* spp.) and crop
41 species including varieties of barley (*Hordeum vulgare*), cotton, wheat (*Triticum aestivum*) and
42 *Brassica* spp (Mahmood et al., 2016) as well as fruit crops in the form of Date Palm (*Phoenix*
43 *dactylifera*) that has high salt tolerance with no visible adverse effects on seedlings (Yaish and
44 Kumar, 2015; Al-Mulla et al., 2013; Alrasbi et al., 2010). Pomegranate (*Punica granatum L.*) is
45 another fruit crop of moderate to high salt tolerance. Through regulating growth form and nutrient
46 balancing, it can maintain water content, chlorophyll fluorescence and enzyme activity at normal
47 levels (Ibrahim, 2016; Okhovatian-Ardakani et al., 2010).

1 In India and elsewhere, tree species including *Prosopis juliflora*, *Dalbergia sissoo*, *Eucalyptus*
2 *tereticornis* have been used to revegetate saline land. Certain biofuel crops in the form of *Ricinus*
3 *communis* (Abideen et al., 2014), *Euphorbia antisyphilitica* (Dagar et al., 2014), *Karelinia caspia*
4 (Akinshina et al., 2016) and *Salicornia* spp. (Sanandiyana and Siddhanta, 2014) are grown in saline
5 areas, and *Panicum turgidum* (Koyro et al., 2013) and *Leptochloa fusca* (Akhter et al., 2003) have
6 been grown as fodder crop on degraded soils with brackish water. In China, intense efforts are being
7 made on the use of halophytes (Sakai et al., 2012; Wang et al., 2018). These examples reveal that
8 there is great scope still use saline areas in a productive manner through the utilisation of halophytes.
9 The most productive species often have yields equivalent to conventional crops, at salinity levels
10 matching even that of sea water.

11 **3.6.2. Socio-economic Responses**

12 Socio-economic and policy responses are often crucial in enhancing the adoption of SLM practices
13 (Cordingley et al., 2015; Fleskens and Stringer, 2014; Nyanga et al., 2016) and for assisting
14 agricultural households to diversify their sources of income (Barrett et al., 2017; Shiferaw and Djido,
15 2016). Technology and socio-economic responses are not independent, but continuously interact.

16 **3.6.2.1. Socio-economic Responses for Combating Desertification Under Climate Change**

17 Desertification limits the choice of potential climate change mitigation and adaptation response
18 options by reducing climate change adaptive capacities. Furthermore, many additional factors, for
19 example, a lack of access to markets or insecurity of land tenure, hinder the adoption of SLM. These
20 factors are largely beyond the control of individuals or local communities and require broader policy
21 interventions (3.6.3). Nevertheless, local collective action and indigenous and local knowledge are
22 still crucial to the ability of households to respond to the combined challenge of climate change and
23 desertification. Raising awareness, capacity building and development to promote collective action
24 and indigenous and local knowledge contribute to avoiding, reducing and reversing desertification
25 under changing climate.

26 ***The use of indigenous and local knowledge*** enhances the success of SLM and its ability to address
27 desertification (Altieri and Nicholls, 2017; Engdawork and Bork, 2016). Using indigenous and local
28 knowledge for combating desertification could contribute to climate change adaptation strategies
29 (Belfer et al., 2017; Codjoe et al., 2014; Etchart, 2017; Speranza et al., 2010; Makondo and Thomas,
30 2018; Maldonado et al., 2016; Nyong et al., 2007). There are abundant examples of how indigenous
31 and local knowledge, which are an important part of broader agroecological knowledge (Altieri,
32 2018), have allowed livelihood systems in drylands to be maintained despite environmental
33 constraints. An example is the numerous traditional water harvesting techniques that are used across
34 the drylands to adapt to dry spells and climate change. These include creating planting pits (“zai”,
35 “ngoro”) and micro-basins, contouring hill slopes and terracing (Biazin et al., 2012) (3.6.1).
36 Traditional “ndiva” water harvesting system in Tanzania enables the capture of runoff water from
37 highland areas to downstream community-managed micro-dams for subsequent farm delivery through
38 small scale canal networks (Enfors and Gordon, 2008). A further example are pastoralist communities
39 located in drylands who have developed numerous methods to sustainably manage rangelands.
40 Pastoralist communities in Morocco developed the “agdal” system of seasonally alternating use of
41 rangelands to limit overgrazing (Dominguez, 2014) as well as to manage forests in the Moroccan
42 High Atlas Mountains (Auclair et al., 2011). Across the Arabian Peninsula and North Africa, a
43 rotational grazing system “hema” was historically practiced by the Bedouin communities (Hussein,
44 2011; Louhaichi and Tastad, 2010). The Beni-Amer herders in the Horn of Africa have developed
45 complex livestock breeding and selection systems (Fre, 2018). Although well adapted to resource-
46 sparse dryland environments, traditional practices are currently not able to cope with increased
47 demand for food and environmental changes (Enfors and Gordon, 2008; Engdawork and Bork, 2016).
48 Moreover, there is *robust evidence* documenting the marginalisation or loss of indigenous and local

1 knowledge (Dominguez, 2014; Fernández-Giménez and Fillat Estaque, 2012; Hussein, 2011;
2 Kodirekkala, 2017; Moreno-Calles et al., 2012). Combined use of indigenous and local knowledge
3 and new SLM technologies can contribute to raising resilience to the challenges of climate change and
4 desertification (*high confidence*) (Engdawork and Bork, 2016; Guzman et al., 2018).

5 **Collective action** has the potential to contribute to SLM and climate change adaptation (*medium*
6 *confidence*) (Adger, 2003; Engdawork and Bork, 2016; Eriksen and Lind, 2009; Ostrom, 2009;
7 Rodima-Taylor et al., 2012). Collective action is a result of social capital. Social capital is divided
8 into structural and cognitive forms, structural corresponding to strong networks (including outside
9 one's immediate community) and cognitive encompassing mutual trust and cooperation within
10 communities (van Rijn et al., 2012; Woolcock and Narayan, 2000). Social capital is more important
11 for economic growth in settings with weak formal institutions, and less so in those with strong
12 enforcement of formal institutions (Ahlerup et al., 2009). There are cases throughout the drylands
13 showing that community bylaws and collective action successfully limited land degradation and
14 facilitated SLM (Ajayi et al., 2016; Infante, 2017; Kassie et al., 2013; Nyangena, 2008; Willy and
15 Holm-Müller, 2013; Wossen et al., 2015). However, there are also cases when they did not improve
16 SLM where they were not strictly enforced (Teshome et al., 2016). Collective action for implementing
17 responses to dryland degradation is often hindered by local asymmetric power relations and "elite
18 capture" (Kihui, 2016; Stringer et al., 2007). This illustrates that different levels and types of social
19 capital result in different levels of collective action. In a sample of East, West and southern African
20 countries, structural social capital in the form of access to networks outside one's own community
21 was suggested to stimulate the adoption of agricultural innovations, whereas cognitive social capital,
22 associated with inward-looking community norms of trust and cooperation, was found to have a
23 negative relationship with the adoption of agricultural innovations (van Rijn et al., 2012). The latter is
24 indirectly corroborated by observations of the impact of community-based rangeland management
25 organisations in Mongolia. Although levels of cognitive social capital did not differ between them,
26 communities with strong links to outside networks were able to apply more innovative rangeland
27 management practices in comparison to communities without such links (Ulambayar et al., 2017).

28 **Farmer-led innovations.** Agricultural households are not just passive adopters of externally
29 developed technologies, but are active experimenters and innovators (Reij and Waters-Bayer, 2001;
30 Tambo and Wünscher, 2015; Waters-Bayer et al., 2009). SLM technologies co-generated through
31 direct participation of agricultural households have higher chances of being accepted by them
32 (*medium confidence*) (Bonney et al., 2016; Vente et al., 2016). Usually farmer-driven innovations are
33 more frugal and better adapted to their resource scarcities than externally introduced technologies
34 (Gupta et al., 2016). Farmer-to-farmer sharing of their own innovations and mutual learning positively
35 contribute to higher technology adoption rates (Dey et al., 2017). This innovative ability can be given
36 a new dynamism by combining it with emerging external technologies. For example, emerging low-
37 cost phone applications that are linked to soil and water monitoring sensors can provide farmers with
38 previously inaccessible information and guidance (Cornell et al., 2013; Herrick et al., 2017; McKinley
39 et al., 2017; Steger et al., 2017).

40 Currently, the adoption of SLM practices remains insufficient to address desertification and contribute
41 to climate change adaptation and mitigation more extensively. This is due to the constraints on the use
42 of indigenous and local knowledge and collective action, as well as economic and institutional
43 barriers for SLM adoption (3.1.4.2; 3.6.3; Banadda, 2010; Cordingley et al., 2015; Lokonon and
44 Mbaye, 2018; Mulinge et al., 2016; Wildemeersch et al., 2015). Sustainable development of drylands
45 under these socio-economic and environmental (climate change, desertification) conditions will also
46 depend on the ability of dryland agricultural households to diversify their livelihoods sources
47 (Boserup, 1965; Safriel and Adeel, 2008).

3.6.2.2. Socio-Economic Responses for Economic Diversification

Livelihood diversification through non-farm employment increases the resilience of rural households against desertification and extreme weather events by diversifying their income and consumption (*high confidence*). Moreover, it can provide the funds to invest into SLM (Belay et al., 2017; Bryan et al., 2009; Dumenu and Obeng, 2016; Salik et al., 2017; Shiferaw et al., 2009). Access to non-agricultural employment is especially important for poorer pastoral households as their small herd sizes make them less resilient to drought (Fratkin, 2013; Lybbert et al., 2004). However, access to alternative opportunities is limited in the rural areas of many developing countries, especially for women and marginalised groups who lack education and social networks (Reardon et al., 2008).

Migration is frequently used as an adaptation strategy to environmental change (*medium confidence*). Migration is a form of livelihood diversification and a potential response option to desertification and increasing risk to agricultural livelihoods under climate change (Walther et al., 2002). Migration can be short-term (e.g., seasonal) or long-term, internal within a country or international. There is *medium evidence* showing rural households responding to desertification and droughts through all forms of migration, for example: during the Dust Bowl in the United States in the 1930s (Hornbeck, 2012); during droughts in Burkina Faso in the 2000s (Barbier et al., 2009); in Mexico in the 1990s (Nawrotzki et al., 2016); and by the Aymara people of the semiarid Tarapacá region in Chile between 1820-1970 responding to declines in rainfall and growing demands for labor outside the region (Lima et al., 2016). There is *robust evidence and high agreement* showing that migration decisions are influenced by a complex set of different factors, with desertification and climate change playing relatively lesser roles (Liehr et al., 2016) (3.4.2). Barrios et al. (2006) found that urbanisation in Sub-Saharan Africa was partially influenced by climatic factors during the 1950 to 2000 period, in parallel to liberalisation of internal restrictions on labour movements: with 1% reduction in rainfall associated with 0.45% increase in urbanisation. This migration favoured more industrially-diverse urban areas in Sub-Saharan Africa (Henderson et al., 2017), because they offer more diverse employment opportunities and higher wages. Similar trends were also observed in Iran in response to water scarcity (Madani et al., 2016). However, migration involves some initial investments. For this reason, reductions in agricultural incomes due to climate change or desertification have the potential to decrease out-migration among the poorest agricultural households who become less able to afford migration (Cattaneo and Peri, 2016), thus increasing social inequalities. There is *medium evidence and high agreement* that households with migrant worker members are more resilient against extreme weather events and environmental degradation compared to non-migrant households who are more dependent on agricultural income (Liehr et al., 2016; Salik et al., 2017; Sikder and Higgins, 2017). Remittances from migrant household members potentially contribute to SLM adoptions, however, substantial out-migration was also found to constrain the implementation of labour-intensive land management practices (Chen et al., 2014; Liu et al., 2016a).

3.6.3. Policy Responses

The adoption of SLM practices depends on the compatibility of the technology with prevailing socio-economic and biophysical conditions (Sanz et al., 2017). Globally, it was shown that every USD invested into restoring degraded lands yields social returns, including both provisioning and non-provisioning ecosystem services, in the range of USD 3–6 over a 30-year period (Nkonya et al., 2016a). A similar range of returns from land restoration activities were found in Central Asia (Mirzabaev et al., 2016), Ethiopia (Gebreselassie et al., 2016), India (Mythili and Goedecke, 2016), Kenya (Mulinge et al., 2016), Niger (Moussa et al., 2016) and Senegal (Sow et al., 2016). Despite these relatively high returns, there is *robust evidence* that the adoption of SLM practices remains low (Cordingley et al., 2015; Giger et al., 2015; Lokonon and Mbaye, 2018). Part of the reason for these low adoption rates is that the major share of the returns from SLM are social benefits, namely in the form of non-provisioning ecosystem services (Nkonya et al., 2016a). The adoption of SLM

1 technologies does not always provide implementers with immediate private benefits (Schmidt et al.,
2 2017). High initial investment costs, institutional and governance constraints and a lack of access to
3 technologies and equipment may inhibit their adoption further (Giger et al., 2015; Sanz et al., 2017;
4 Schmidt et al., 2017). However, not all SLM practices have high upfront costs. Analysing the World
5 Overview of Conservation Approaches and Technologies (WOCAT) database, a globally
6 acknowledged reference database for SLM, Giger et al. (2015) found that the upfront costs of SLM
7 technologies ranged from about USD 20 to USD 5000, with the median cost being around USD 500 .
8 Many SLM technologies are profitable within three to 10 years (*medium evidence, high agreement*)
9 (Djanibekov and Khamzina, 2016; Giger et al., 2015; Moussa et al., 2016; Sow et al., 2016). About
10 73% of 363 SLM technologies evaluated were reported to become profitable within three years, while
11 97% were profitable within 10 years (Giger et al., 2015). Similarly, it was shown that social returns
12 from investments in restoring degraded lands will exceed their costs within six years in many settings
13 across drylands (Nkonya et al., 2016a). However, even with affordable upfront costs, market failures
14 in the form of lack of access to credit, input and output markets, and insecure land tenure (3.1.3) result
15 in the lack of adoption of SLM technologies (Moussa et al., 2016). Payments for ecosystem services,
16 subsidies for SLM, encouragement of community collective action can lead to a higher level of
17 adoption of SLM and land restoration activities (*medium confidence*) (Bouma and Wösten, 2016;
18 Lambin et al., 2014; Reed et al., 2015; Schiappacasse et al., 2012; van Zanten et al., 2014; 3.6.3).
19 Enabling policy responses discussed in this section contribute to overcoming these market failures.

20 Many socio-economic factors shaping individual responses to desertification typically operate at
21 larger scales. Individual households and communities do not exercise control over these factors, such
22 as land tenure insecurity, lack of property rights, lack of access to markets, availability of rural
23 advisory services, and agricultural price distortions. These factors are shaped by national government
24 policies and international markets. As in the case with socio-economic responses, policy responses are
25 classified below in two ways: those which seek to combat desertification under changing climate; and
26 those which seek to provide alternative livelihood sources through economic diversification. These
27 options are mutually complementary and contribute to all the three hierarchical elements of the Land
28 Degradation Neutrality (LDN) framework, namely, avoiding, reducing and reversing land degradation
29 (Cowie et al., 2018; Orr et al., 2017; 4.8.5; Table 7.2; 7.4.5). Enabling policy environment is a critical
30 element for the achievement of LDN (Chasek et al., 2019). Implementation of LDN policies can
31 contribute to climate change adaptation and mitigation (*high confidence*) (3.6.1, 3.7.2).

3.6.3.1. Policy Responses towards Combating Desertification under Climate Change

32 Policy responses to combat desertification take numerous forms (Marques et al., 2016). Below we
33 discuss major policy responses consistently highlighted in the literature in connection with SLM and
34 climate change, because these response options were found to strengthen adaptation capacities and to
35 contribute to climate change mitigation. They include improving market access, empowering women,
36 expanding access to agricultural advisory services, strengthening land tenure security, payments for
37 ecosystem services, decentralised natural resource management, investing into research and
38 monitoring of desertification and dust storms, and investing into modern renewable energy sources.

39
40 ***Policies aiming at improving market access***, that is the ability to access output and input markets at
41 lower costs, help farmers and livestock producers earn more profit from their produce. Increased
42 profits both motivate and enable them to invest more in SLM. Higher access to input, output and
43 credit markets was consistently found as a major factor in the adoption of SLM practices in a wide
44 number of settings across the drylands (*medium confidence*) (Aw-Hassan et al., 2016; Gebreselassie et
45 al., 2016; Mythili and Goedecke, 2016; Nkonya and Anderson, 2015; Sow et al., 2016). Lack of
46 access to credit limits adjustments and agricultural responses to the impacts of desertification under
47 changing climate, with long-term consequences on the livelihoods and incomes, as was shown for the
48 case of the American Dust Bowl during 1930s (Hornbeck, 2012). Government policies aimed at

1 improving market access usually involve constructing and upgrading rural-urban transportation
2 infrastructure and agricultural value chains, such as investments into construction of local markets,
3 abattoirs and cold storage warehouses, as well as post-harvest processing facilities (Mcpeak et al.,
4 2006). However, besides infrastructural constraints, providing improved access often involves
5 relieving institutional constraints to market access (Little, 2010), such as improved coordination of
6 cross-border food safety and veterinary regulations (Ait Hou et al., 2015; Keiichiro et al., 2015;
7 Mcpeak et al., 2006; Unnevehr, 2015), and availability and access to market information systems
8 (Bobojonov et al., 2016; Christy et al., 2014; Nakasone et al., 2014).

9 **Women's empowerment.** A greater emphasis on understanding gender-specific differences over land-
10 use and land management practices as an entry point can make land restoration projects more
11 successful (*medium confidence*) (Broeckhoven and Cliquet, 2015; Carr and Thompson, 2014;
12 Catacutan and Villamor, 2016; Dah-gbeto and Villamor, 2016). In relation to representation and
13 authority to make decisions in land management and governance, women's participation remains
14 lacking particularly in the dryland regions. Thus, ensuring women's rights means accepting women as
15 equal members of the community and citizens of the state (Nelson et al., 2015). This includes
16 equitable access of women to resources (including extension services), networks, and markets. In
17 areas where socio-cultural norms and practices devalue women and undermine their participation,
18 actions for empowering women will require changes in customary norms, recognition of women's
19 (land) rights in government policies and programmes to assure that their interests are better
20 represented (1.4.2; Cross-Chapter Box 11: Gender, Chapter 7). In addition, several novel concepts are
21 recently applied for an in-depth understanding of gender in relation to science-policy interface.
22 Among these are the concepts of intersectionality, i.e. how social dimensions of identity and gender
23 are bound up in systems of power and social institution (Thompson-Hall et al., 2016), bounded
24 rationality for gendered decision making, related to incomplete information interacting with limits to
25 human cognition leading to judgement errors or objectively poor decision making (Villamor and van
26 Noordwijk, 2016), anticipatory learning for preparing for possible contingencies and consideration of
27 long-term alternatives (Dah-gbeto and Villamor, 2016) and systematic leverage points for
28 interventions that produce, mark, and entrench gender inequality within communities (Manlosa et al.,
29 2018), which all aim to improve gender equality within agro-ecological landscapes through a systems
30 approach.

31 **Education and expanding access to agricultural services.** Providing access to information about
32 SLM practices facilitates their adoption (*medium confidence*) (Kassie et al., 2015; Nkonya et al.,
33 2015; Nyanga et al., 2016). Moreover, improving the knowledge of climate change, capacity building
34 and development in rural areas can help strengthen climate change adaptive capacities (Berman et al.,
35 2012; Chen et al., 2018; Descheemaeker et al., 2018; Popp et al., 2009; Tambo, 2016; Yaro et al.,
36 2015). Agricultural initiatives to improve the adaptive capacities of vulnerable populations were more
37 successful when they were conducted through reorganised social institutions and improved
38 communication, e.g. in Mozambique (Osbaahr et al., 2008). Improved communication and education
39 could be facilitated by wider use of new information and communication technologies (Peters et al.,
40 2015). Investments into education were associated with higher adoption of soil conservation
41 measures, e.g. in Tanzania (Tenge et al., 2004). Bryan et al. (2009) found that access to information
42 was the prominent facilitator of climate change adaptation in Ethiopia. However, resource constraints
43 of agricultural services, and disconnects between agricultural policy and climate policy can hinder the
44 dissemination of climate smart agricultural technologies (Morton, 2017). Lack of knowledge was also
45 found to be a significant barrier to implementation of soil rehabilitation programmes in the
46 Mediterranean region (Reichardt, 2010). Agricultural services will be able to facilitate SLM best
47 when they also serve as platforms for sharing indigenous and local knowledge and farmer innovations
48 (Mapfumo et al., 2016). Participatory research initiatives conducted jointly with farmers have higher
49 chances of resulting in technology adoption (Bonney et al., 2016; Rusike et al., 2006; Vente et al.,

1 2016). Moreover, rural advisory services are often more successful in disseminating technological
2 innovations when they adopt commodity/value chain approaches, remain open to engagement in input
3 supply, make use of new opportunities presented by information and communication technologies
4 (ICTs), facilitate mutual learning between multiple stakeholders (Morton, 2017), and organise science
5 and SLM information in a location-specific manner for use in education and extension (Bestelmeyer
6 et al., 2017).

7 **Strengthening land tenure security.** Strengthening land tenure security is a major factor contributing
8 to the adoption of soil conservation measures in croplands (*high confidence*) (Bambio and Bouayad
9 Agha, 2018; Higgins et al., 2018; Holden and Ghebru, 2016; Paltasingh, 2018; Rao et al., 2016;
10 Robinson et al., 2018) , thus, contributing to climate change adaptation and mitigation. Moreover,
11 land tenure security can lead to more investment in trees (Deininger and Jin, 2006; Etongo et al.,
12 2015). Land tenure recognition policies were found to lead to higher agricultural productivity and
13 incomes, although with inter-regional variations, requiring an improved understanding of overlapping
14 formal and informal land tenure rights (Lawry et al., 2017). For example, secure land tenure increased
15 investments into SLM practices in Ghana, however, without affecting farm productivity (Abdulai et
16 al., 2011). Secure land tenure, especially for communally managed lands, helps reduce arbitrary
17 appropriations of land for large scale commercial farms (Aha and Ayitey, 2017; Baumgartner, 2017;
18 Dell'Angelo et al., 2017). In contrast, privatisation of rangeland tenures in Botswana and Kenya led to
19 the loss of communal grazing lands and actually increased rangeland degradation (Basupi et al., 2017;
20 Kihui, 2016) as pastoralists needed to graze livestock on now smaller communal pastures. Since food
21 insecurity in drylands is strongly affected by climate risks, there is *robust evidence and high*
22 *agreement* that resilience to climate risks is higher with flexible tenure for allowing mobility for
23 pastoralist communities, and not fragmenting their areas of movement (Behnke, 1994; Holden and
24 Ghebru, 2016; Liao et al., 2017; Turner et al., 2016; Wario et al., 2016). More research is needed on
25 the optimal tenure mix, including low-cost land certification, redistribution reforms, market-assisted
26 reforms and gender-responsive reforms, as well as collective forms of land tenure such as communal
27 land tenure and cooperative land tenure (see 7.6.5 for a broader discussion of land tenure security
28 under climate change).

29 **Payment for ecosystem services (PES)** provide incentives for land restoration and SLM (*medium*
30 *confidence*) (Lambin et al., 2014; Li et al., 2018; Reed et al., 2015; Schiappacasse et al., 2012).
31 Several studies illustrate that social cost of desertification are larger than its private cost (Costanza et
32 al., 2014; Nkonya et al., 2016a). Therefore, although SLM can generate public goods in the form of
33 provisioning ecosystem services, individual land custodians underinvest in SLM as they are unable to
34 reap these benefits fully. Payment for ecosystem services provides a mechanism through which some
35 of these benefits can be transferred to land users, thereby stimulating further investment in SLM. The
36 effectiveness of PES schemes depends on land tenure security and appropriate design taking into
37 account specific local conditions (Börner et al., 2017). However, PES has not worked well in
38 countries with fragile institutions (Karsenty and Ongolo, 2012). Equity and justice in distributing the
39 payments for ecosystem services were found to be key for the success of the PES programmes in
40 Yunnan, China (He and Sikor, 2015). Yet, when reviewing the performance of PES programmes in
41 the tropics, Calvet-Mir et al. (2015), found that they are generally effective in terms of environmental
42 outcomes, despite being sometimes unfair in terms of payment distribution. It is suggested that the
43 implementation of PES will be improved through decentralised approaches giving local communities
44 a larger role in the decision making process (He and Lang, 2015).

45 **Empowering local communities for decentralised natural resource management.** Local institutions
46 often play a vital role in implementing SLM initiatives and climate change adaptation (*high*
47 *confidence*) (Gibson et al., 2005; Smucker et al., 2015). Pastoralists involved in community-based
48 natural resource management in Mongolia had greater capacity to adapt to extreme winter frosts

1 resulting in less damage to their livestock (Fernandez-Gimenez et al., 2015). Decreasing the power
2 and role of traditional community institutions, due to top-down public policies, resulted in lower
3 success rates in community-based programmes focused on rangeland management in Dirre, Ethiopia
4 (Abdu and Robinson, 2017). Decentralised governance was found to lead to improved management in
5 forested landscapes (Dressler et al., 2010; Ostrom and Nagendra, 2006). However, there are also cases
6 when local elites were placed in control, decentralised natural resource management negatively
7 impacted the livelihoods of the poorer and marginalised community members due to reduced access
8 to natural resources (Andersson and Ostrom, 2008; Cullman, 2015; Dressler et al., 2010). The success
9 of decentralised natural resource management initiatives depends on increased participation and
10 empowerment of diverse set of community members, not only local leaders and elites, in the design
11 and management of local resource management institutions (Kadirbeyoglu and Özertan, 2015;
12 Umutoni et al., 2016), while considering the interactions between actors and institutions at different
13 levels of governance (Andersson and Ostrom, 2008; Carlisle and Gruby, 2017; McCord et al., 2017).
14 An example of such programmes where local communities played a major role in land restoration and
15 rehabilitation activities is the cooperative project on “The National Afforestation and Erosion Control
16 Mobilization Action Plan” in Turkey, initiated by the Turkish Ministry of Agriculture and Forestry
17 (Çalışkan and Boydak, 2017), with the investment of USD 1.8 billion between 2008 and 2012. The
18 project mobilised local communities in cooperation with public institutions, municipalities, and non-
19 governmental organisations, to implement afforestation, rehabilitation and erosion control measures,
20 resulting in the afforestation and reforestation of 1.5 M ha (Yurtoglu, 2015). Moreover, some 1.75 M
21 ha of degraded forest and 37880 ha of degraded rangelands were rehabilitated. Finally, the project
22 provided employment opportunities for 300,000 rural residents for six months every year, combining
23 land restoration and rehabilitation activities with measures to promote socio-economic development in
24 rural areas (Çalışkan and Boydak, 2017).

25 ***Investing in research and development.*** Desertification has received substantial research attention
26 over recent decades (Turner et al., 2007). There is also a growing research interest on climate change
27 adaptation and mitigation interventions that help address desertification (Grainger, 2009). Agricultural
28 research on SLM practices has generated a significant number of new innovations and technologies
29 that increase crop yields without degrading the land, while contributing to climate change adaptation
30 and mitigation (3.6.1). There is *robust evidence* that such technologies help improve the food security
31 of smallholder dryland farming households (Harris and Orr, 2014, 6.3.5). Strengthening research on
32 desertification is of high importance not only to meet SDGs but also effectively manage ecosystems
33 based on solid scientific knowledge. More investment in research institutes and training the younger
34 generation of researchers is needed for addressing the combined challenges of desertification and
35 climate change (Akhtar-Schuster et al., 2011; Verstraete et al., 2011). This includes improved
36 knowledge management systems that allow stakeholders to work in a coordinated manner by
37 enhancing timely, targeted and contextualised information sharing (Chasek et al., 2011). Knowledge
38 and flow of knowledge on desertification is currently highly fragmented, constraining effectiveness of
39 those engaged in assessing and monitoring the phenomenon at various levels (Reed et al., 2011).
40 Improved knowledge and data exchange and sharing increase the effectiveness of efforts to address
41 desertification (*high confidence*).

42 ***Developing modern renewable energy sources.*** Transitioning to renewable energy resources
43 contributes to reducing desertification by lowering reliance on traditional biomass in dryland regions
44 (*medium confidence*). Populations in most developing countries continue to rely on traditional
45 biomass, including fuelwood, crop straws and livestock manure, for a major share of their energy
46 needs, with the highest dependence in Sub-Saharan Africa (Amugune et al., 2017; IEA, 2013). Use of
47 biomass for energy, mostly fuelwood (especially as charcoal), was associated with deforestation in
48 some dryland areas (Iiyama et al., 2014; Mekuria et al., 2018; Neufeldt et al., 2015; Zulu, 2010),
49 while in some other areas there was no link between fuelwood collection and deforestation (Simon

1 and Peterson, 2018; Swemmer et al., 2018; Twine and Holdo, 2016). Moreover, the use of traditional
2 biomass as a source of energy was found to have negative health effects through indoor air pollution
3 (de la Sota et al., 2018; Lim and Seow, 2012), while also being associated with lower female labor
4 force participation (Burke and Dundas, 2015). Jiang et al., (2014) indicated that providing improved
5 access to alternative energy sources such as solar energy and biogas could help reduce the use of
6 fuelwood in south-western China, thus alleviating the spread of rocky desertification. The conversion
7 of degraded lands into cultivation of biofuel crops will affect soil C dynamics (Albanito et al., 2016;
8 Nair et al., 2011; Cross-Chapter Box 7: Bioenergy and BECCS, Chapter 6). The use of biogas slurry
9 as soil amendment or fertiliser can increase soil C (Galvez et al., 2012; Negash et al., 2017). Large-
10 scale installation of wind and solar farms in the Sahara desert was projected to create a positive
11 climate feedback through increased surface friction and reduced albedo, doubling precipitation over
12 the neighbouring Sahel region with resulting increases in vegetation (Li et al., 2018). Transition to
13 renewable energy sources in high-income countries in dryland areas primarily contributes to reducing
14 greenhouse gas emissions and mitigating climate change, with some other co-benefits such as
15 diversification of energy sources (Bang, 2010), while the impacts on desertification are less evident.
16 The use of renewable energy has been proposed as an important mitigation option in dryland areas as
17 well (El-Fadel et al., 2003). Transitions to renewable energy are being promoted by governments
18 across drylands (Cancino-Solórzano et al., 2016; Hong et al., 2013; Sen and Ganguly, 2017) including
19 in fossil-fuel rich countries (Farnoosh et al., 2014; Dehkordi et al., 2017; Stambouli et al., 2012;
20 Vidadili et al., 2017), despite important social, political and technical barriers to expanding renewable
21 energy production (Afsharzade et al., 2016; Baker et al., 2014; Elum and Momodu, 2017; Karatayev
22 et al., 2016). Improving the social awareness about the benefits of transitioning to renewable energy
23 resources and access to hydro-energy, solar and wind energy contributes to their improved adoption
24 (Aliyu et al., 2017; Katikiro, 2016).

25 ***Developing and strengthening climate services relevant for desertification.*** Climate services provide
26 climate, drought and desertification-related information in a way that assists decision making by
27 individuals and organisations. For monitoring desertification, integration of biogeophysical (climate,
28 soil, ecological factors, biodiversity) and socio-economic aspects (use of natural resources by local
29 population) provides a basis for better vulnerability prediction and assessment (OSS, 2012; Vogt et
30 al., 2011). Examples of relevant services include: drought monitoring and early warning systems often
31 implemented by national climate and meteorological services but also encompassing regional and
32 global systems (Pozzi et al., 2013); and the Sand and Dust Storm Warning Advisory and Assessment
33 System (SDS-WAS), created by WMO in 2007, in partnership with the World Health Organization
34 (WHO) and the United Nations Environment Program (UNEP). Currently, there is also a lack of
35 ecological monitoring in arid and semi-arid regions to study surface winds, dust and sandstorms, and
36 their impacts on ecosystems and human health (Bergametti et al., 2018; Marticorena et al., 2010).
37 Reliable and timely climate services, relevant to desertification, can aid the development of
38 appropriate adaptation and mitigation options reducing the impact of desertification under changing
39 climate on human and natural systems (*high confidence*) (Beegum et al., 2016; Beegum et al., 2018;
40 Cornet, 2012; Haase et al., 2018; Sergeant, Moynahan, & Johnson, 2012).

41 ***3.6.3.2. Policy Responses Supporting Economic Diversification***

42 Despite policy responses for combating desertification, climate change, growing food demands, as
43 well as the need to reduce poverty and strengthen food security, will put strong pressures on the land
44 (Cherlet et al., 2018; 6.1.4; 7.2.2). Sustainable development of drylands and their resilience to
45 combined challenges of desertification and climate change will thus also depend on the ability of
46 governments to promote policies for economic diversification within agriculture and in non-
47 agricultural sectors in order make dryland areas less vulnerable to desertification and climate change.

1 **Investing into irrigation.** Investments into expanding irrigation in dryland areas can help increase the
2 resilience of agricultural production to climate change, improve labour productivity and boost
3 production and income revenue from agriculture and livestock sectors (Geerts and Raes, 2009;
4 Olayide et al., 2016; Oweis and Hachum, 2006). This is particularly true for Sub-Saharan Africa,
5 where currently only 6% of the cultivated areas are irrigated (Nkonya et al., 2016b). While renewable
6 groundwater resources could help increase the share of irrigated land to 20.5%-48.6% of croplands in
7 the region (Altchenko and Villholth, 2015). On the other hand, over-extraction of groundwaters,
8 mainly for irrigating crops, is becoming an important environmental problem in many dryland areas
9 (Cherlet et al., 2018), requiring careful design and planning of irrigation expansion schemes and use
10 of water efficient irrigation methods (Bjornlund, van Rooyen, and Stirzaker, 2017; Woodhouse et al.,
11 2017). For example, in Saudi Arabia, improving the efficiency of water management, e.g. through the
12 development of aquifers, water recycling and rainwater harvesting is part of policy actions to combat
13 desertification (Bazza, et al., 2018; Kingdom of Saudi Arabia, 2016). The expansion of irrigation to
14 riverine areas, crucial for dry season grazing of livestock, needs to consider the loss of income from
15 pastoral activities, which is not always lower than income from irrigated crop production (Behnke and
16 Kerven, 2013). Irrigation development could be combined with the deployment of clean energy
17 technologies in economically viable ways (Chandel et al., 2015). For example, solar-powered drip
18 irrigation was found to increase household agricultural incomes in Benin (Burney et al., 2010). The
19 sustainability of irrigation schemes based on solar-powered extraction of groundwaters depends on
20 measures to avoid over-abstraction of groundwater resources and associated negative environmental
21 impacts (Closas and Rap, 2017).

22 **Expanding agricultural commercialisation.** Faster poverty rate reduction and economic growth
23 enhancement is realised when countries transition into the production of non-staple, high value
24 commodities and manage to build a robust agro-industry sector (Barrett et al., 2017). Ogotu and Qaim
25 (2019) found that agricultural commercialisation increased incomes and decreased multidimensional
26 poverty in Kenya. Similar findings were earlier reported by Muriithi and Matz (2015) for
27 commercialisation of vegetables in Kenya. Commercialisation of rice production was found to have
28 increased smallholder welfare in Nigeria (Awotide et al., 2016). Agricultural commercialisation
29 contributed to improved household food security in Malawi, Tanzania and Uganda (Carletto et al.,
30 2017). However, such a transition did not improve farmers' livelihoods in all cases (Reardon et al.,
31 2009). High value cash crop/animal production can be bolstered by wide-scale use of technologies,
32 for example, mechanisation, application of inorganic fertilisers, crop protection and animal health
33 products. Market oriented crop/animal production facilitates social and economic progress with labour
34 increasingly shifting out of agriculture into non-agricultural sectors (Cour, 2001). Modernised
35 farming, improved access to inputs, credit and technologies enhances competitiveness in local and
36 international markets (Reardon et al., 2009).

37 **Facilitating structural transformations** in rural economies implies that the development of non-
38 agricultural sectors encourages the movement of labour from land-based livelihoods, vulnerable to
39 desertification and climate change, to non-agricultural activities (Haggblade et al., 2010). The
40 movement of labour from agriculture to non-agricultural sectors is determined by relative labour
41 productivities in these sectors (Shiferaw and Djido, 2016). Given already high underemployment in
42 the farm sector, increasing labour productivity in the non-farm sector was found as the main driver of
43 labour movements from farm sector to non-farm sector (Shiferaw and Djido, 2016). More investments
44 into education can facilitate this process (Headey et al., 2014). However, in some contexts, such as
45 pastoralist communities in Xinjiang, China, income diversification was not found to improve the
46 welfare of pastoral households (Liao et al., 2015). Economic transformations also occur through
47 urbanisation, involving the shift of labour from rural areas into gainful employment in urban areas
48 (Jedwab and Vollrath, 2015). The larger share of world population will be living in urban centres in
49 the 21st century and this will require innovative means of agricultural production with minimum

1 ecological footprint and less dependence on fossil fuels (Revi and Rosenzweig, 2013), while
2 addressing the demand of cities (see 4.9.1 for discussion on urban green infrastructure). Although
3 there is some evidence of urbanisation leading to the loss of indigenous and local ecological
4 knowledge, however, indigenous and local knowledge systems are constantly evolving, and are also
5 getting integrated into urban environments (Júnior et al., 2016; Reyes-García et al., 2013; van Andel
6 and Carvalheiro, 2013). Urban areas are attracting an increasing number of rural residents across the
7 developing world (Angel et al., 2011; Cour, 2001; Dahiya, 2012). Urban development contributes to
8 expedited agricultural commercialisation by providing market outlet for cash and high value crop and
9 livestock products. At the same time, urbanisation also poses numerous challenges in the form of
10 rapid urban sprawl and pressures on infrastructure and public services, unemployment and associated
11 social risks, which have considerable implications on climate change adaptive capacities (Bulkeley,
12 2013; Garschagen and Romero-Lankao, 2015).

13

14

15 **Cross-Chapter Box 5: Policy Responses to Drought**

16 Alisher Mirzabaev (Germany/Uzbekistan), Margot Hurlbert (Canada), Muhammad Mohsin Iqbal
17 (Pakistan), Joyce Kimutai (Kenya), Lennart Olsson (Sweden), Fasil Tena (Ethiopia), Murat Türkeş
18 (Turkey)

19

20 Drought is a highly complex natural hazard (for floods, see Box 7.2). It is difficult to precisely
21 identify its start and end. It is usually slow and gradual (Wilhite and Pulwarty, 2017), but sometimes
22 can evolve rapidly (Ford and Labosier, 2017; Mo and Lettenmaier, 2015). It is context-dependent, but
23 its impacts are diffuse, both direct and indirect, short-term and long-term (Few and Tebboth, 2018;
24 Wilhite and Pulwarty, 2017). Following the Synthesis Report (SYR) of the IPCC Fifth Assessment
25 Report (AR5), drought is defined here as “a period of abnormally dry weather long enough to cause a
26 serious hydrological imbalance” (Mach et al., 2014). Although drought is considered abnormal
27 relative to the water availability under the mean climatic characteristics, it is also a recurrent element
28 of any climate, not only in drylands, but also in humid areas (Cook et al., 2014b; Seneviratne and
29 Ciais, 2017; Spinoni et al., 2019; Türkeş, 1999; Wilhite et al., 2014). Climate change is projected to
30 increase the intensity or frequency of droughts in some regions across the world (for detailed
31 assessment see 2.2, and IPCC Special Report on Global Warming of 1.5°C (Hoegh-Guldberg et al.,
32 2018, Chapter 3)). Droughts often amplify the effects of unsustainable land management practices,
33 especially in drylands, leading to land degradation (Cook et al., 2009; Hornbeck, 2012). Especially in
34 the context of climate change, the recurrent nature of droughts requires pro-actively planned policy
35 instruments both to be well-prepared to respond to droughts when they occur and also undertake ex
36 ante actions to mitigate their impacts by strengthening the societal resilience against droughts (Gerber
37 and Mirzabaev, 2017).

38 Droughts are among the costliest of natural hazards (*robust evidence, high agreement*). According to
39 the International Disaster Database (EM-DAT), droughts affected more than 1.1 billion people
40 between 1994-2013, with the recorded global economic damage of USD 787 billion (CRED, 2015),
41 corresponding to an average of USD 41.4 billion per year. Drought losses in the agricultural sector
42 alone in the developing countries were estimated to equal USD 29 billion between 2005-2015 (FAO,
43 2018). Usually, these estimates capture only direct and on-site costs of droughts. However, droughts
44 have also wide-ranging indirect and off-site impacts, which are seldom quantified. These indirect
45 impacts are both biophysical and socio-economic, with the poor households and communities being
46 particularly exposed to them (Winsemius et al., 2018). Droughts affect not only water quantity, but
47 also water quality (Mosley, 2014). The costs of these water quality impacts are yet to be adequately

1 quantified. Socio-economic indirect impacts of droughts are related to food insecurity, poverty,
2 lowered health and displacement (Gray and Mueller, 2012; Johnstone and Mazo, 2011; Linke et al.,
3 2015; Lohmann and Lechtenfeld, 2015; Maystadt and Ecker, 2014; Yusa et al., 2015 see also 3.4.2.9,
4 Box 5.5), which are difficult to quantify comprehensively. Research is required for developing
5 methodologies that could allow for more comprehensive assessment of these indirect drought costs.
6 Such methodologies require the collection of highly granular data, which is currently lacking in many
7 countries due to high costs of data collection. However, the opportunities provided by remotely
8 sensed data and novel analytical methods based on big data and artificial intelligence, including use of
9 citizen science for data collection, could help in reducing these gaps.

10 There are three broad (and sometimes overlapping) policy approaches for responding to droughts
11 (also see 7.4.8). These approaches are often pursued simultaneously by many governments. Firstly,
12 responding to drought when it occurs by providing direct drought relief, known as crisis management.
13 Crisis management is also the costliest among policy approaches to droughts because it often
14 incentivises the continuation of activities vulnerable to droughts (Botterill and Hayes, 2012; Gerber
15 and Mirzabaev, 2017).

16 The second approach involves development of drought preparedness plans, which coordinate the
17 policies for providing relief measures when droughts occur. For example, combining resources to
18 respond to droughts at regional level in Sub-Saharan Africa was found more cost-effective than
19 separate individual country drought relief funding (Clarke and Hill, 2013). Effective drought
20 preparedness plans require well-coordinated and integrated government actions - a key lesson learnt
21 from 2015-2017 drought response in Cape Town, South Africa (Visser, 2018). Reliable, relevant and
22 timely climate and weather information helps respond to droughts appropriately (Sivakumar and
23 Ndiang'ui, 2007). Improved knowledge and integration of weather and climate information can be
24 achieved by strengthening drought early warning systems at different scales (Verbist et al., 2016).
25 Every USD invested into strengthening hydro-meteorological and early warning services in
26 developing countries was found to yield between USD 4 to 35 (Hallegatte, 2012). Improved access
27 and coverage by drought insurance, including index insurance, can help alleviate the impacts of
28 droughts on livelihoods (Guerrero-Baena et al., 2019; Kath et al., 2019; Osgood et al., 2018; Ruiz et
29 al., 2015; Tadesse et al., 2015).

30 The third category of responses to droughts involves drought risk mitigation. Drought risk mitigation
31 is a set of proactive measures, policies and management activities aimed at reducing the future
32 impacts of droughts (Vicente-Serrano et al., 2012). For example, policies aimed at improving water
33 use efficiency in different sectors of the economy, especially in agriculture and industry, or public
34 advocacy campaigns raising societal awareness and bringing about behavioural change to reduce
35 wasteful water consumption in the residential sector are among such drought risk mitigation policies
36 (Tsakiris, 2017). Public outreach and monitoring of communicable diseases, air and water quality
37 were found useful for reducing health impacts of droughts (Yusa et al., 2015). The evidence from
38 household responses to drought in Cape Town, South Africa, between 2015-2017, suggests that media
39 coverage and social media could play a decisive role in changing water consumption behaviour, even
40 more so than official water consumption restrictions (Booyesen et al., 2019). Drought risk mitigation
41 approaches are less costly than providing drought relief after the occurrence of droughts. To illustrate,
42 Harou et al. (2010) found that establishment of water markets in California considerably reduced
43 drought costs. Application of water saving technologies reduced drought costs in Iran by USD 282
44 million (Salami et al., 2009). Booker et al. (2005) calculated that interregional trade in water could
45 reduce drought costs by 20–30% in the Rio Grande basin, USA. Increasing rainfall variability under
46 climate change can make the forms of index insurance based on rainfall less efficient (Kath et al.,
47 2019). A number of diverse water property instruments, including instruments allowing water
48 transfer, together with the technological and institutional ability to adjust water allocation, can

1 improve timely adjustment to droughts (Hurlbert, 2018). Supply-side water management providing for
2 proportionate reductions in water delivery prevents the important climate change adaptation option of
3 managing water according to need or demand (Hurlbert and Mussetta, 2016). Exclusive use of a water
4 market to govern water allocation similarly prevents the recognition of the human right to water at
5 times of drought (Hurlbert, 2018). Policies aiming to secure land tenure, to expand access to markets,
6 agricultural advisory services and effective climate services, as well as to create off-farm employment
7 opportunities can facilitate the adoption of drought risk mitigation practices (Alam, 2015; Kusunose
8 and Lybbert, 2014), increasing the resilience to climate change (3.6.3), while also contributing to
9 SLM (3.6.3, 4.8.1, Table 5.7).

10 The excessive burden of drought relief funding on public budgets is already leading to a paradigm
11 shift towards proactive drought risk mitigation instead of reactive drought relief measures (Verner et
12 al., 2018; Wilhite, 2016). Climate change will reinforce the need for such proactive drought risk
13 mitigation approaches. Policies for drought risk mitigation that are already needed now will be even
14 more relevant under higher warming levels (Jerneck and Olsson, 2008; McLeman, 2013; Wilhite et
15 al., 2014). Overall, there is *high confidence* that responding to droughts through ex post drought relief
16 measures is less efficient compared to ex ante investments into drought risk mitigation, particularly
17 under climate change.

18 3.6.4. Limits to Adaptation, Maladaptation, and Barriers for Mitigation

19 Chapter 16 in the Fifth Assessment Report of IPCC (Klein et al., 2015) discusses the existence of soft
20 and hard limits to adaptation, highlighting that values and perspectives of involved agents are relevant
21 to identify limits (4.8.5.1, 7.4.9). In that sense, adaptation limits vary from place to place and are
22 difficult to generalise (Barnett et al., 2015; Dow et al., 2013; Klein et al., 2015). Currently, there is a
23 lack of knowledge on adaptation limits and potential maladaptation to combined effects of climate
24 change and desertification (see 4.8.6 in Chapter 4 for discussion on resilience, thresholds, and
25 irreversible land degradation also relevant for desertification). However, the potential for residual
26 risks and maladaptive outcomes is high (*high confidence*). Some examples of residual risks are
27 illustrated below (those risks which remain after adaptation efforts were taken, irrespective whether
28 they are tolerable or not, tolerability being a subjective concept). Although SLM measures can help
29 lessen the effects of droughts, they cannot fully prevent water stress in crops and resulting lower
30 yields (Eekhout and de Vente, 2019). Moreover, although in many cases SLM measures can help
31 reduce and reverse desertification, there would be still short-term losses in land productivity.
32 Irreversible forms of land degradation (e.g. loss of topsoil, severe gully erosion) can lead to the
33 complete loss of land productivity. Even when solutions are available, their costs could be prohibitive
34 presenting the limits to adaptation (Dixon et al., 2013). If warming in dryland areas surpasses human
35 thermal physiological thresholds (Klein et al., 2015; Waha et al., 2013), adaptation could eventually
36 fail (Kamali et al., 2018). Catastrophic shifts in ecosystem functions and services, e.g. coastal erosion
37 (4.9.8; Chen et al., 2015; Schneider and Kéfi, 2016), and economic factors can also result in
38 adaptation failure (Evans et al., 2015). Despite the availability of numerous options that contribute to
39 combating desertification, climate change adaptation and mitigation, there are also chances of
40 maladaptive actions (*medium confidence*) (Glossary). Some activities favouring agricultural
41 intensification in dryland areas can become maladaptive due to their negative impacts on the
42 environment (*medium confidence*). Agricultural expansion to meet food demands can come through
43 deforestation and consequent diminution of C sinks (Godfray and Garnett, 2014; Stringer et al., 2012).
44 Agricultural insurance programs encouraging higher agricultural productivity and measures for
45 agricultural intensification can result in detrimental environmental outcomes in some settings
46 (Guodaar et al., 2019; Müller et al., 2017; Table 6.12). Development of more drought-tolerant crop
47 varieties is considered as a strategy for adaptation to shortening rainy season, but this can also lead to
48 a loss of local varieties (Al Hamndou and Requier-Desjardins, 2008). Livelihood diversification to

1 collecting and selling firewood and charcoal production can exacerbate deforestation (Antwi-Agyei et
2 al., 2018). Avoiding maladaptive outcomes can often contribute both to reducing the risks from
3 climate change and combating desertification (Antwi-Agyei et al., 2018). Avoiding, reducing and
4 reversing desertification would enhance soil fertility, increase C storage in soils and biomass, thus
5 reducing C emissions from soils to the atmosphere (3.7.2; Cross-Chapter Box 2: Implications of large-
6 scale conversion from non-forest to forest land, Chapter 1). In specific locations, there may be barriers
7 for some of these activities. For example, afforestation and reforestation programs can contribute to
8 reducing sand storms and increasing C sinks in dryland regions (3.6.1, 3.7.2) (Chu et al., 2019).
9 However, implementing agroforestry measures in arid locations can be constrained by lack of water
10 (Apuri et al., 2018), leading to a trade-off between soil C sequestration and other water uses (Cao et
11 al., 2018).

13 **3.7. Hotspots and Case Studies**

14 The challenges of desertification and climate change in dryland areas across the world often have very
15 location-specific characteristics. The five case studies in this section present rich experiences and
16 lessons learnt on: 1) soil erosion, 2) afforestation and reforestation through “green walls”, 3) invasive
17 plant species, 4) oases in hyper-arid areas, and 5) integrated watershed management. Although it is
18 impossible to cover all hotspots of desertification and on the ground actions from all dryland areas,
19 these case studies present a more focused assessment of these five issues that emerged as salient in the
20 group discussions and several rounds of review of this chapter. The choice of these case studies was
21 also motivated by the desire to capture a wide diversity of dryland settings.

22 **3.7.1. Climate Change and Soil Erosion**

23 ***3.7.1.1. Soil Erosion under Changing Climate in Drylands***

24 Soil erosion is a major form of desertification occurring in varying degrees in all dryland areas across
25 the world (3.2), with negative effects on dryland ecosystems (3.4). Climate change is projected to
26 increase soil erosion potential in some dryland areas through more frequent heavy rainfall events and
27 rainfall variability than currently (see Section 3.5.2 for more detailed assessment, (Achite and Ouillon,
28 2007; Megnounif and Ghenim, 2016; Vachtman et al., 2013; Zhang and Nearing, 2005). There are
29 numerous soil conservation measures that can help reduce soil erosion (3.6.1). Such soil management
30 measures include afforestation and reforestation activities, rehabilitation of degraded forests, erosion
31 control measures, prevention of overgrazing, diversification of crop rotations, and improvement in
32 irrigation techniques, especially in sloping areas (Anache et al., 2018; ÇEMGM, 2017; Li and Fang,
33 2016; Poesen, 2018; Ziadat and Taimeh, 2013). Effective measures for soil conservation can also use
34 spatial patterns of plant cover to reduce sediment connectivity, and the relationships between
35 hillslopes and sediment transfer in eroded channels (García-Ruiz et al., 2017). The following three
36 examples present lessons learnt from the soil erosion problems and measures to address them in
37 different settings of Chile, Turkey and the Central Asian countries.

38 ***3.7.1.2. No-Till Practices for Reducing Soil Erosion in Central Chile***

39 Soil erosion by water is an important problem in Chile. National assessments conducted in 1979,
40 which examined 46% of the continental surface of the country, concluded that very high levels of soil
41 erosion affected 36% of the territory. The degree of soil erosion increases from south to north. The
42 leading locations in Chile are the region of Coquimbo with 84% of eroded soils (Lat 29°S, Semiarid
43 climate), the region of Valparaíso with 57% of eroded souls (Lat 33° S, Mediterranean climate) and
44 the region of O'Higging with 37% of eroded soils (Lat 34°S Mediterranean climate). The most
45 important drivers of soil erosion are soil, slope, climate erosivity (i.e., precipitation, intensity, duration
46 and frequency) due to a highly concentrated rainy season, and vegetation structure and cover. In the
47 region of Coquimbo, goat and sheep overgrazing have aggravated the situation (CIREN, 2010).

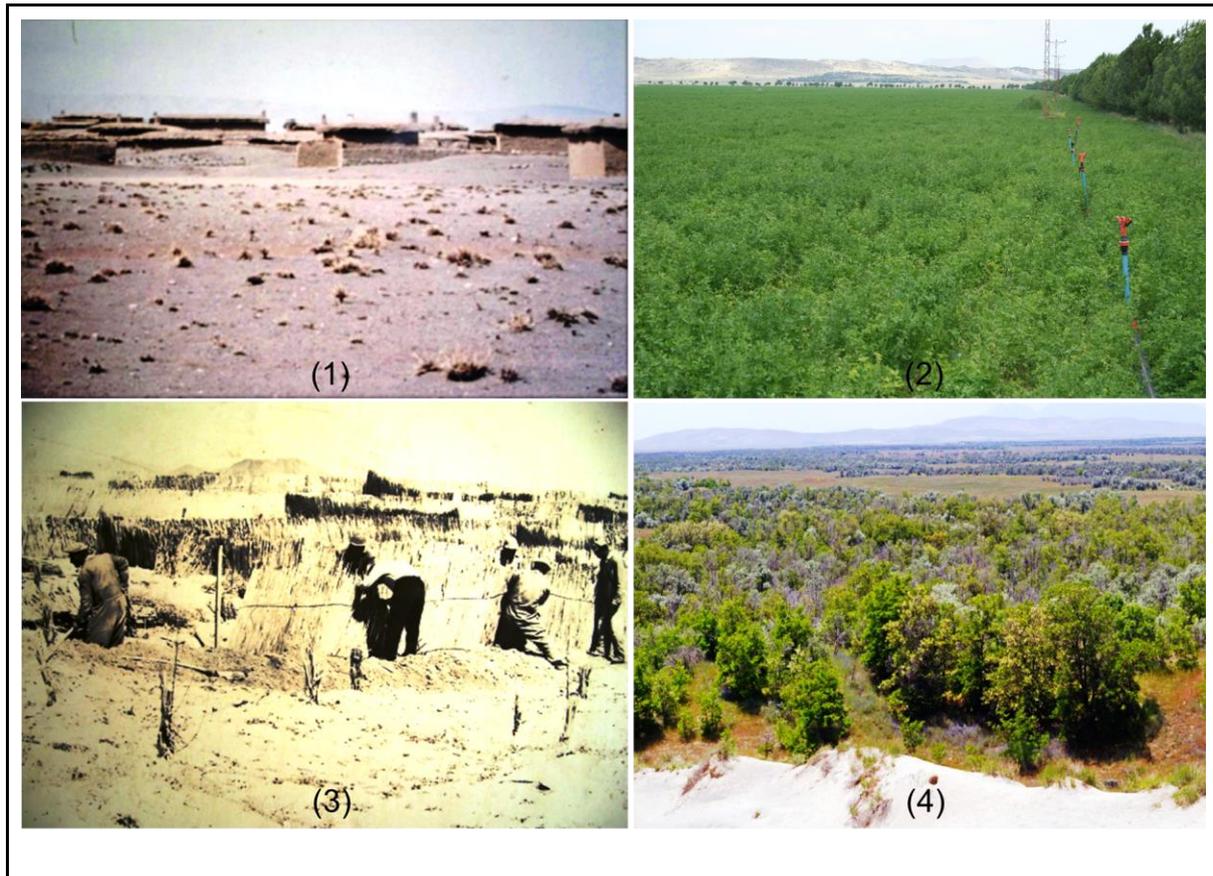
1 Erosion rates reach up to 100 t ha⁻¹ annually, having increased substantially over the last 50 years
2 (Ellies, 2000). About 10.4% of central Chile exhibits high erosion rates (greater than 1.1 t ha⁻¹
3 annually) (Bonilla et al., 2010).

4 Over the last few decades there has been an increasing interest in the development of no-till (also
5 called zero tillage) technologies to minimise soil disturbance, reduce the combustion of fossil fuels
6 and increase soil organic matter. No-till in conjunction with the adoption of strategic cover crops have
7 positively impacted soil biology with increases in soil organic matter. Early evaluations by Crovetto,
8 (1998) showed that no-till application (after seven years) had doubled the biological activity
9 indicators compared to traditional farming and even surpassed those found in pasture (grown for the
10 previous 15 years). Besides erosion control, additional benefits are an increase of water holding
11 capacity and reduction in bulk density. Currently, the above no-till farm experiment has lasted for 40
12 years and continues to report benefits to soil health and sustainable production (Reicosky and
13 Crovetto, 2014). The influence of this iconic farm has resulted in the adoption of soil conservation
14 practices and specially no-till in dryland areas of the Mediterranean climate region of central Chile
15 (Martínez et al., 2011). Currently, it has been estimated that the area under no-till farming in Chile
16 varies between 0.13 and 0.2 M ha (Acevedo and Silva, 2003).

17 ***3.7.1.3. Combating Wind Erosion and Deflation in Turkey: The Greening Desert of*** 18 ***Karapınar***

19 In Turkey, the amount of sediment recently released through erosion into seas was estimated to be 168
20 Mt yr⁻¹, which is considerably lower than the 500 Mt yr⁻¹ that was estimated to be lost in the 1970s
21 (ÇEMGM, 2017). The decrease in erosion rates is attributed to an increase in spatial extent of forests,
22 rehabilitation of degraded forests, erosion control, prevention of overgrazing, and improvement in
23 irrigation technologies. Soil conservation measures conducted in the Karapınar district, Turkey,
24 exemplify these activities. The district is characterised by a semi-arid climate and annual average
25 precipitation of 250–300 mm (Türkeş, 2003; Türkeş and Tatlı, 2011). In areas where vegetation was
26 overgrazed or inappropriately tilled, the surface soil horizon was removed through erosion processes
27 resulting in the creation of large drifting dunes that threatened settlements around Karapınar
28 (Groneman, 1968). Such dune movement had begun to affect the Karapınar settlement in 1956
29 (Kantarıcı et al., 2011). Consequently, by early 1960s, Karapınar town and nearby villages were
30 confronted with the danger of abandonment due to out-migration in early 1960s (Figure 3.11-1). The
31 reasons for increasing wind erosion in the Karapınar district can be summarised as follows: sandy
32 material was mobilised following drying of the lake; hot and semi-arid climate conditions;
33 overgrazing and use of pasture plants for fuel; excessive tillage; and strong prevailing winds.

34



1 **Figure 3.11 (1) A general view of a nearby village of Karapınar town in early 1960s (Çarkacı, 1999). (2) A**
 2 **view of the Karapınar wind erosion area in 2013 (Photograph: Murat Türkeş, 17.06.2013). (3)**
 3 **Construction of cane screens in early 1960s in order to decrease speed of the wind and prevent**
 4 **movement of the sand accumulations and dunes, which was one of the physical measures during the**
 5 **prevention and mitigation period (Çarkacı, 1999). (4) A view of mix vegetation in most of the Karapınar**
 6 **wind erosion area in 2013, the main tree species of which were selected for afforestation with respect to**
 7 **their resistance to the arid continental climate conditions along with a warm/hot temperature regime over**
 8 **the district (Photograph: Murat Türkeş, 17.06.2013)**

9
 10 Restoration and mitigation strategies were initiated in 1959 and today, 4300 ha of land have been
 11 restored (Akay and Yildirim, 2010) (Figure 3.11-2), using specific measures: (1) Physical measures:
 12 construction of cane screens to decrease wind speed and prevent sand movement (Figure 3.11); (2)
 13 Restoration of cover: increasing grass cover between screens using seeds collected from local pastures
 14 or the cultivation of rye (*Secale* sp.) and wheat grass (*Agropyron elongatum*) that are known to grow
 15 in arid and hot conditions; (3) Afforestation: saplings obtained from nursery gardens were planted and
 16 grown between these screens. Main tree species selected were oleaster (*Eleagnus* sp.), acacia (*Robinia*
 17 *pseudeaccacia*), ash (*Fraxinus* sp.), elm (*Ulmus* sp.) and maple (*Acer* sp.) (Figure 3.11-4). Economic
 18 growth occurred after controlling erosion and new tree nurseries have been established with modern
 19 irrigation. Potential negative consequences through the excessive use of water can be mitigated
 20 through engagement with local stakeholders and transdisciplinary learning processes, as well as by
 21 restoring the traditional land uses in the semi-arid Konya closed basin (Akça et al., 2016).

22 **3.7.1.4. Soil Erosion in Central Asia under Changing Climate**

23
 24 Soil erosion is widely acknowledged to be a major form of degradation of Central Asian drylands,
 25 affecting considerable share of croplands and rangelands. However, up-to-date information on the
 26 actual extent of eroded soils at the regional or country level is not available. The estimates compiled
 27 by Pender et al. (2009), based on the Central Asian Countries Initiative for Land Management

1 (CACILM), indicate that about 0.8 M ha of the irrigated croplands were subject to high degree of soil
2 erosion in Uzbekistan. In Turkmenistan, soil erosion was indicated to be occurring in about 0.7 M ha
3 of irrigated land. In Kyrgyzstan, out of 1 M ha irrigated land in the foothill zones, 0.76 M ha were
4 subject to soil erosion by water, leading to losses in crop yields of 20-60% in these eroded soils.
5 About 0.65 M ha of arable land were prone to soil erosion by wind (Mavlyanova et al., 2017).
6 Soil erosion is widespread in rainfed and irrigated areas in Kazakhstan (Saparov, 2014). About 5 M ha
7 of rainfed croplands were subject to high levels of soil erosion (Pender et al., 2009). Soil erosion by
8 water was indicated to be a major concern in sloping areas in Tajikistan (Pender et al., 2009).

9 The major causes of soil erosion in Central Asia are related to human factors, primarily excessive
10 water use in irrigated areas (Gupta et al., 2009), deep ploughing and lack of maintenance of vegetative
11 cover in rainfed areas (Suleimenov et al., 2014), and overgrazing in rangelands (Mirzabaev et al.,
12 2016). Lack of good maintenance of watering infrastructure for migratory livestock grazing and
13 fragmentation of livestock herds led to overgrazing near villages, increasing the soil erosion by wind
14 (Alimaev et al., 2008). Overgrazing in the rangeland areas of the region (e.g. particularly in
15 Kyzylkum) contributes to dust storms, coming primarily from Ustyurt Plateau, desertified areas of
16 Amudarya and Syrdarya rivers' deltas, dried seabed of the Aral Sea (now called Aralkum), and the
17 Caspian Sea (Issanova and Abuduwaili, 2017; Xi and Sokolik, 2015). Xi and Sokolik (2015)
18 estimated that total dust emissions in Central Asia were 255.6 Mt in 2001, representing 10-17% of the
19 global total.

20 Central Asia is one of the regions highly exposed to climate change, with warming levels projected to
21 be higher than the global mean (Hoegh-Guldberg et al., 2018), leading to more heat extremes (Reyer
22 et al., 2017). There is no clear trend in precipitation extremes, with some potential for moderate rise in
23 occurrence of droughts. The diminution of glaciers is projected to continue in the Pamir and Tian
24 Shan mountain ranges, a major source of surface waters along with seasonal snowmelt. Glacier
25 melting will increase the hazards from moraine-dammed glacial lakes and spring floods (Reyer et al.,
26 2017). Increased intensity of spring floods creates favourable conditions for higher soil erosion by
27 water especially in the sloping areas in Kyrgyzstan and Tajikistan. The continuation of some of the
28 current unsustainable cropland and rangeland management practices may lead to elevated rates of soil
29 erosion particularly in those parts of the region where climate change projections point to increases in
30 floods (Kyrgyzstan, Tajikistan) or increases in droughts (Turkmenistan, Uzbekistan) (Hijioka et al.,
31 2014). Increasing water use to compensate for higher evapotranspiration due to growing temperatures
32 and heat waves could increase soil erosion by water in the irrigated zones, especially sloping areas
33 and crop fields with uneven land levelling (Bekchanov et al., 2010). The desiccation of the Aral Sea
34 resulted in hotter and drier regional microclimate, adding to the growing wind erosion in adjacent
35 deltaic areas and deserts (Kust, 1999).

36 There are numerous sustainable land and water management practices available in the region for
37 reducing soil erosion (Abdullaev et al., 2007; Gupta et al., 2009; Kust et al., 2014; Nurbekov et al.,
38 2016). These include: improved land levelling and more efficient irrigation methods such as drip,
39 sprinkler and alternate furrow irrigation (Gupta et al., 2009); conservation agriculture practices,
40 including no-till methods and maintenance of crop residues as mulch in the rainfed and irrigated areas
41 (Kienzler et al., 2012; Pulatov et al., 2012); rotational grazing; institutional arrangements for pooling
42 livestock for long-distance mobile grazing; reconstruction of watering infrastructure along the
43 livestock migratory routes (Han et al., 2016; Mirzabaev et al., 2016); afforesting degraded marginal
44 lands (Djanibekov and Khamzina, 2016; Khamzina et al., 2009; Khamzina et al., 2016); integrated
45 water resource management (Dukhovny et al., 2013; Kazbekov et al., 2009), planting salt and drought
46 tolerant halophytic plants as windbreaks in sandy rangelands (Akinshina et al., 2016; Qadir et al.,
47 2009; Toderich et al., 2009; Toderich et al., 2008), and potentially the dried seabed of the former Aral
48 Sea (Breckle, 2013). The adoption of enabling policies, such as those discussed in Section 3.6.3, can

1 facilitate the adoption of these sustainable land and water management practices in Central Asia (*high*
2 *confidence*) (Aw-Hassan et al., 2016; Bekchanov et al., 2016; Bobojonov et al., 2013; Djanibekov et
3 al., 2016; Hamidov et al., 2016; Mirzabaev et al., 2016).

4 **3.7.2. Green Walls and Green Dams**

5 This case study evaluates the experiences of measures and actions implemented to combat soil
6 erosion, decrease dust storms, and to adapt to and mitigate climate change under the Green Wall and
7 Green Dam programmes in East Asia (e.g., China) and Africa (e.g., Algeria, Sahara and the Sahel
8 region). These measures have also been implemented in other countries, such as Mongolia (Do &
9 Kang, 2014; Lin et al., 2009), Turkey (Yurtoglu, 2015; Çalışkan and Boydak, 2017) and Iran
10 (Amiraslani and Dragovich, 2011), and are increasingly considered as part of many national and
11 international initiatives to combat desertification (Goffner et al., 2019; Cross-Chapter Box 2, chapter
12 1). Afforestation and reforestation programs can contribute to reducing sand storms and increasing C
13 sinks in dryland regions (*high confidence*). On the other hand, Green Wall and Green Dam
14 programmes also decrease the albedo and hence increase the surface absorption of radiation,
15 increasing the surface temperature. The net effect will largely depend on the balance between these
16 and will vary from place to place depending on many factors.

17 **3.7.2.1. The Experiences of Combating Desertification in China**

18 Arid and semiarid areas of China, including north-eastern, northern and north-western regions, cover
19 an area of more than 509 M ha, with annual rainfall of below 450 mm. Over the past several centuries,
20 more than 60% of the areas in arid and semiarid regions were used as pastoral and agricultural lands.
21 The coupled impacts of past climate change and human activity have caused desertification and dust
22 storms to become a serious problem in the region (Xu et al., 2010). In 1958, the Chinese government
23 recognised that desertification and dust storms jeopardised livelihoods of nearly 200 million people,
24 and afforestation programmes for combating desertification have been initiated since 1978. China is
25 committed to go beyond the Land Degradation Neutrality objective as indicated by the following
26 programmes that have been implemented. The Chinese Government began the Three North's Forest
27 Shelterbelt programme in Northeast China, North China, and Northwest China, with the goal to
28 combat desertification and to control dust storms by improving forest cover in arid and semiarid
29 regions. The project is implemented in three stages (1978–2000, 2001–2020, and 2021–2050). In
30 addition, the Chinese government launched Beijing and Tianjin Sandstorm Source Treatment Project
31 (2001–2010), Returning Farmlands to Forest Project (2003–present), Returning Grazing Land to
32 Grassland Project (2003–present) to combat desertification, and for adaptation and mitigation of
33 climate change (State Forestry Administration of China, 2015; Tao, 2014; Wang et al., 2013).

34 The results of the fifth period monitoring (2010–2014) showed: (1) Compared with 2009, the area of
35 degraded land decreased by 12,120 km² over a five-year period; (2) In 2014, the average coverage of
36 vegetation in the sand area was 18.33%, an increase of 0.7% compared with 17.63% in 2009, and the
37 C sequestration increased by 8.5%; (3) Compared with 2009, the amount of wind erosion decreased
38 by 33%, the average annual occurrence of sandstorms decreased by 20.3% in 2014; (4) As of 2014,
39 203,700 km² of degraded land were effectively managed, accounting for 38.4% of the 530,000 km² of
40 manageable desertified land; (5) The restoration of degraded land has created an annual output of
41 53.63 M tonnes of fresh and dried fruits, accounting for 33.9% of the total national annual output of
42 fresh and dried fruits (State Forestry Administration of China, 2015). This has become an important
43 pillar for economic development and a high priority for peasants as a method to eradicate poverty
44 (State Forestry Administration of China, 2015).

45 Stable investment mechanisms for combating desertification have been established along with tax
46 relief policies and financial support policies for guiding the country in its fight against desertification.
47 The investments in scientific and technological innovation for combating desertification have been
48 improved, the technologies for vegetation restoration under drought conditions have been developed,

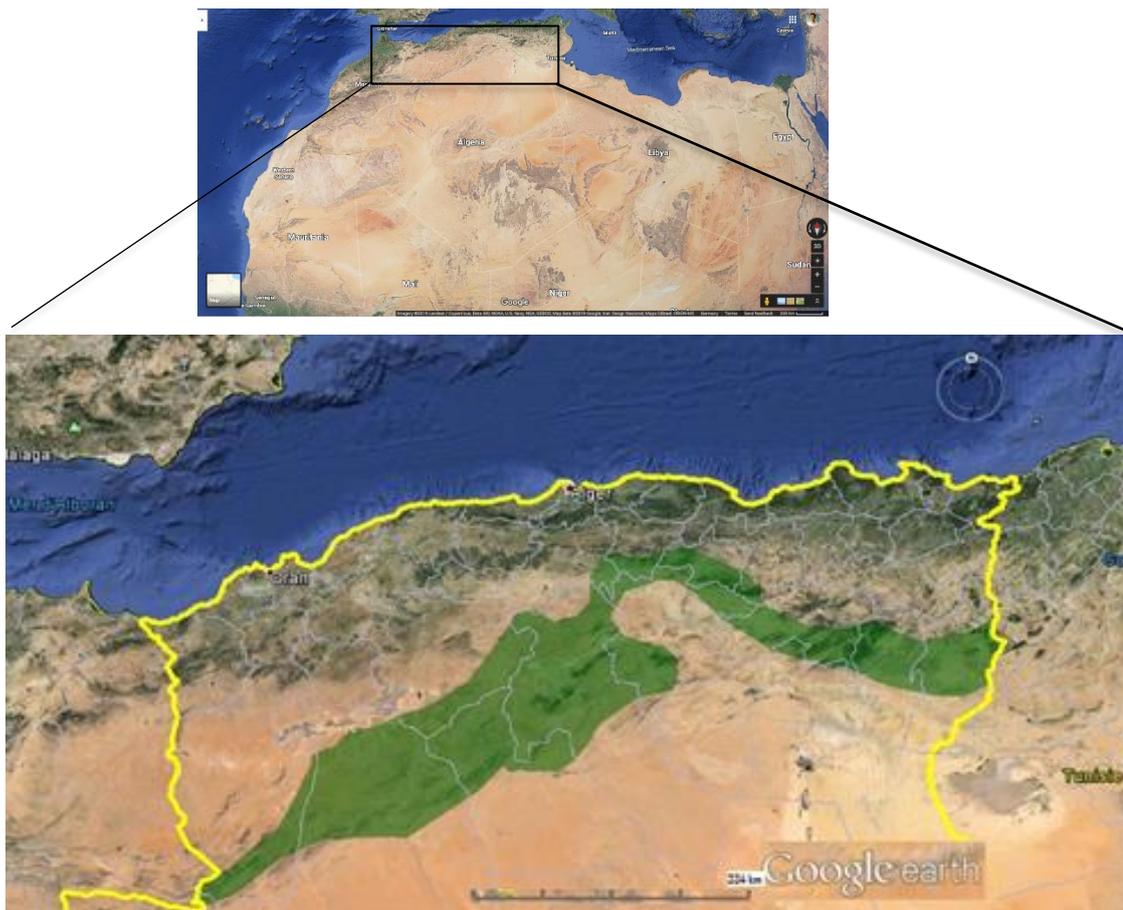
1 the popularisation and application of new technologies has been accelerated, and the training of
2 technicians for farmers and herdsman has been strengthened. To improve the monitoring capability
3 and technical level of desertification, the monitoring network system has been strengthened, and the
4 popularisation and application of modern technologies are intensified (e.g., information and remote
5 sensing) (Wu et al., 2015). Special laws on combating desertification have been decreed by the
6 government. The provincial government responsibilities for desertification prevention and controlling
7 objectives and laws have been strictly implemented.

8 Many studies showed that the projects generally played an active role in combating desertification and
9 fighting against dust storms in China over the past several decades (*high confidence*) (Cao et al.,
10 2018; State Forestry Administration of China, 2015; Wang et al., 2013; Wang et al., 2014; Yang et al.
11 2013). At the beginning of the project, some problems appeared in some places due to lack of enough
12 knowledge and experience (*low confidence*) (Jiang, 2016; Wang et al., 2010). For example, some tree
13 species selected were not well suited to local soil and climatic conditions (Zhu et al., 2007), and there
14 was an inadequate consideration of the limitation of the amount of effective water on the carrying
15 capacity of trees in some arid regions (Dai, 2011; Feng et al., 2016; 3.6.4). In addition, at the
16 beginning of the project, there was an inadequate consideration of the effects of climate change on
17 combating desertification (Feng et al., 2015; Tan and Li, 2015). Indeed, climate change and human
18 activities over past years have influenced the desertification and dust storm control effects in China
19 (Feng et al., 2015; Wang et al., 2009; Tan and Li, 2015), and future climate change will bring new
20 challenges for combating desertification in China (Wang et al., 2017; Yin et al., 2015; Xu et al.,
21 2019). In particular, the desertification risk in China will be enhanced at 2°C compared to 1.5°C
22 global temperature rise (Ma et al., 2018). Adapting desertification control to climate change involves:
23 improving the adaptation capacity to climate change for afforestation and grassland management by
24 executing SLM practices; optimising the agricultural and animal husbandry structure; and using big
25 data to fulfil the water resources regulation (Zhang and Huisinsh, 2018). In particular, improving
26 scientific and technological supports in desertification control is crucial for adaptation to climate
27 change and combating desertification, including protecting vegetation in desertification-prone lands
28 by planting indigenous plant species, facilitating natural restoration of vegetation to conserve
29 biodiversity, employing artificial rain or snow, water saving irrigation and water storage technologies
30 (Jin et al., 2014; Yang et al., 2013).

31 32 **3.7.2.2. The Green Dam in Algeria**

33 After independence in 1962, the Algerian government initiated measures to replant forests destroyed
34 by the war and the steppes affected by desertification among its top priorities (Belaaz, 2003). In 1972,
35 the government invested in the “Green Dam” (“Barrage Vert”) project. This was the first significant
36 experiment to combat desertification, influence the local climate and decrease the aridity by restoring
37 a barrier of trees. The Green Dam extends across arid and semi-arid zones between the isohyets 300
38 and 200 mm. It is a 3 M ha band of plantation running from east to west (Figure 3.12). It is over 1,200
39 km long (from the Algerian-Moroccan border to the Algerian-Tunisian border) and has an average
40 width of about 20 km. The soils in the area are shallow, low in organic matter and susceptible to
41 erosion. The main objectives of the project were to conserve natural resources, improve the living
42 conditions of local residents and avoid their exodus to urban areas. During the first four decades
43 (1970–2000) the success rate was low (42%) due to lack of participation by the local population and
44 the choice of species (Bensaid, 1995).

1



2

3 **Figure 3.12 Localisation of the Green Dam in Algeria (Saifi et al., 2015). Note: The green coloured band**
 4 **represents the location of the Green Dam; the yellow band delineates the national border of Algeria.**
 5 **Source: GoogleEarth**

6

7 The Green Dam did not have the desired effects. Despite tree planting efforts, desertification
 8 intensified on the steppes, especially in south-western Algeria due to the prolonged drought during the
 9 1980s. Rainfall declined from 18% to 27%, and the dry season has increased by two months in the last
 10 century (Belala et al., 2018). Livestock numbers in the Green Dam regions, mainly sheep, have grown
 11 exponentially, leading to severe overgrazing, causing trampling and soil compaction, which greatly
 12 increased the risk of erosion. Wind erosion, very prevalent in the region, is due to climatic conditions
 13 and the strong anthropogenic action that reduced the vegetation cover. The action of the wind carries
 14 fine particles such as sands and clays and leaves on the soil surface a lag gravel pavement, which is
 15 unproductive. Water erosion is largely due to torrential rains in the form of severe thunderstorms that
 16 disintegrate the bare soil surface from raindrop impact (Achite et al., 2016). The detached soil and
 17 nutrients are transported offsite via runoff resulting in loss of fertility and water holding capacity. The
 18 risk of and severity of water erosion is a function of human land use activities that increase soil loss
 19 through removal of vegetative cover. The National Soil Sensitivity to Erosion Map (Salamani et al.,
 20 2012) shows that more than 3 M ha of land in the steppe provinces are currently experiencing intense
 21 wind activity (Houyou et al., 2016) and are areas at particular risk of soil erosion. Mostephaoui et al.
 22 (2013), estimates that each year there is a loss of 7 t h^{-1} of soils due to erosion. Nearly 0.6 M ha of
 23 land in the steppe zone are fully degraded without the possibility of biological recovery.

24

25 To combat the effects of erosion and desertification, the government has planned to relaunch the
 26 rehabilitation of the Green Dam by incorporating new concepts related to sustainable development,
 27 and adaptation to climate change. The experience of previous years has led to integrated rangeland

1 management, improved tree and fodder shrub plantations and the development of water conservation
2 techniques. Reforestation is carried out using several species, including fruit trees, to increase and
3 diversify the sources of income of the population.

4
5 The evaluation of the Green Dam from 1972 to 2015 (Merdas et al., 2015) shows that 0.3 M ha of
6 forest plantation have been planted, which represents 10% of the project area. Estimates of the success
7 rate of reforestation vary considerably between 30% and 75%, depending on the region. Through
8 demonstration, the Green Dam has inspired several African nations to build a Great Green Wall to
9 combat land degradation, mitigate climate change effects, loss of biodiversity and poverty in a region
10 that stretches from Senegal to Djibouti (Sahara and Sahel Observatory (OSS), 2016).

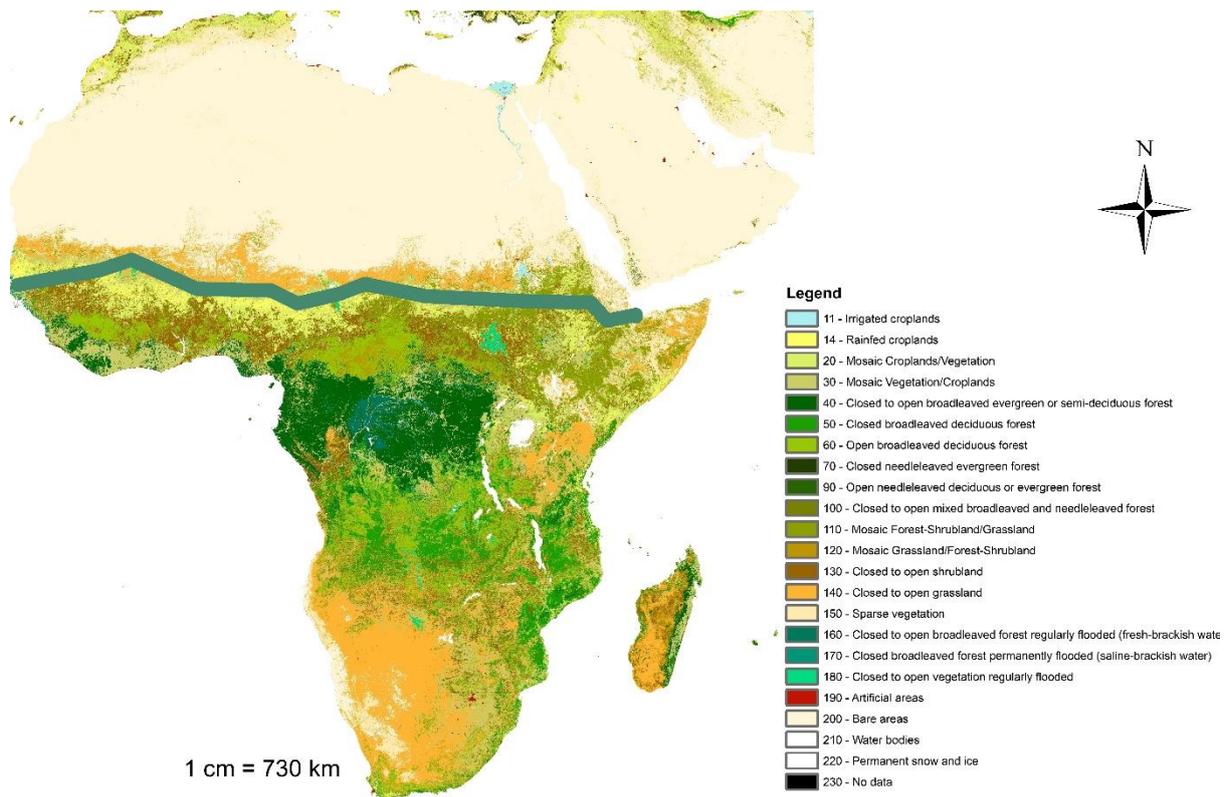
11 12 **3.7.2.3. The Great Green Wall of the Sahara and the Sahel Initiative**

13 The Great Green Wall is an initiative of the Heads of State and Government of the Sahelo-Saharan
14 countries to mitigate and adapt to climate change, and to improve the food security of the Sahel and
15 Saharan peoples (Sacande, 2018; M'Bow, 2017). Launched in 2007, this regional project aims to
16 restore Africa's degraded arid landscapes, reduce the loss of biodiversity and support local
17 communities to sustainable use of forests and rangelands. The Great Green Wall focuses on
18 establishing plantations and neighbouring projects covering a distance of 7,775 km from Senegal on
19 the Atlantic coast to Eritrea on the Red Sea coast, with a width of 15 km (Figure 3.13). The wall
20 passes through Djibouti, Eritrea, Ethiopia, Sudan, Chad, Niger, Nigeria, Mali, Burkina Faso and
21 Mauritania and Senegal.

22 The choice of woody and herbaceous species that will be used to restore degraded ecosystems is
23 based on biophysical and socio-economic criteria, including socio-economic value (food, pastoral,
24 commercial, energetic, medicinal, cultural); ecological importance (C sequestration, soil cover, water
25 infiltration) and species that are resilient to climate change and variability. The Pan-African Agency
26 of the Great Green Wall (PAGGW) was created in 2010 under the auspices of the African Union and
27 CEN-SAD to manage the project. The initiative is implemented at the level of each country by a
28 national structure. A monitoring and evaluation system has been defined, allowing nations to measure
29 outcomes and to propose the necessary adjustments.

30 In the past, reforestation programs in the arid regions of the Sahel and North Africa that have been
31 undertaken to stop desertification were poorly studied and cost a lot of money without significant
32 success (Benjaminsen and Hiernaux, 2019). Today, countries have changed their strategies and opted
33 for rural development projects that can be more easily funded. Examples of scalable practices for land
34 restoration: Managing water bodies for livestock and crop production, promoting fodder trees
35 reducing runoff (Mbow, 2017).

36 The implementation of the initiative has already started in several countries. For example, the FAO's
37 Action Against Desertification project was restoring 18000 hectares of land in 2018 through planting
38 native tree species in Burkina Faso, Ethiopia, the Gambia, Niger, Nigeria and Senegal (Sacande,
39 2018). Berrahmouni et al. (2016) estimated that 166 M ha can be restored in the Sahel, requiring the
40 restoration of 10 M ha per year to achieve Land Degradation Neutrality targets by 2030. Despite this
41 early implementation actions on the ground, the achievement of the planned targets is questionable
42 and challenging without significant additional funding.



1
2
3 **Figure 3.13 The Great Green Wall of the Sahara and the Sahel.**

4 **Source for the data layer: This dataset is an extract from the GlobCover 2009 land cover map, covering**
 5 **Africa and the Arabian Peninsula. The GlobCover 2009 land cover map is derived by an automatic and**
 6 **regionally-tuned classification of a time series of global MERIS (MEdium Resolution Imaging**
 7 **Spectrometer) FR mosaics for the year 2009. The global land cover map counts 22 land cover classes**
 8 **defined with the United Nations (UN) Land Cover Classification System (LCCS).**
 9

10 3.7.3. Invasive Plant Species

11 3.7.3.1. Introduction

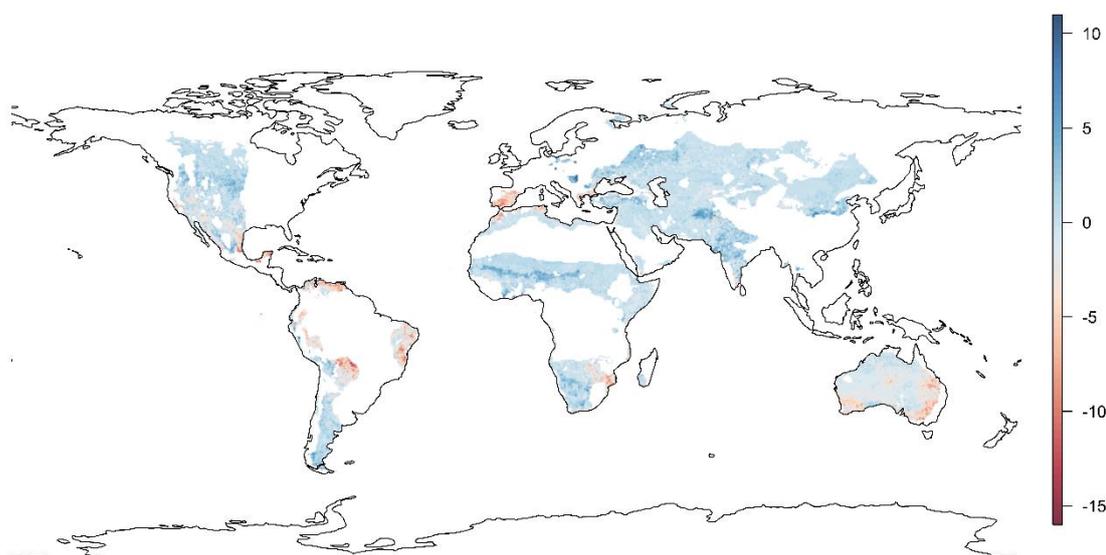
12 The spread of invasive plants can be exacerbated by climate change (Bradley et al., 2010; Davis et al.,
 13 2000). In general, it is expected that the distribution of invasive plant species with high tolerance to
 14 drought or high temperatures may increase under most climate change scenarios (*medium to high*
 15 *confidence*; Bradley et al., 2010; Settele et al., 2014; Scasta et al., 2015). Invasive plants are
 16 considered a major risk to native biodiversity and can disturb the nutrient dynamics and water balance
 17 in affected ecosystems (Ehrenfeld, 2003). Compared to more humid regions, the number of species
 18 that succeed in invading dryland areas is low (Bradley et al., 2012), yet they have a considerable
 19 impact on biodiversity and ecosystem services (Le Maitre et al., 2015; 2011; Newton et al., 2011).
 20 Moreover, human activities in dryland areas are responsible for creating new invasion opportunities
 21 (Safriel et al., 2005).

22 Current drivers of species introductions include expanding global trade and travel, land degradation
 23 and changes in climate (Chytrý et al., 2012; Richardson et al., 2011; Seebens et al., 2018). For
 24 example, Davis et al. (2000) suggests that high rainfall variability promotes the success of alien plant
 25 species - as reported for semiarid grasslands and Mediterranean-type ecosystems (Cassidy et al., 2004;
 26 Reynolds et al., 2004; Sala et al., 2006). Furthermore, Panda et al. (2018) demonstrated that many
 27 invasive species could withstand elevated temperature and moisture scarcity caused by climate

1 change. Dukes et al. (2011) observed that the invasive plant yellow-star thistle (*Centaurea solstitialis*)
2 grew six time larger under elevated atmospheric CO₂ expected in future climate change scenarios.

3 Climate change is *likely* going to aggravate the problem as existing species continue to spread
4 unabated and other species develop invasive characteristics (Hellmann et al., 2008). Although the
5 effects of climate change on invasive species distributions have been relatively well explored, the
6 greater impact on ecosystems is less well understood (Bradley et al., 2010; Eldridge et al., 2011).

7 Due to the time lag between the initial release of invasive species and their impact, the consequence
8 of invasions is not immediately detected and may only be noticed centuries after introduction (Rouget
9 et al., 2016). Climate change and invading species may act in concert (Bellard et al., 2013; Hellmann
10 et al., 2008; Seebens et al., 2015). For example, invasion often changes the size and structure of fuel
11 loads, which can lead to an increase in the frequency and intensity of fire (Evans et al., 2015). In areas
12 where the climate is becoming warmer, an increase in the likelihood of suitable weather conditions for
13 fire may promote invasive species, which in turn may lead to further desertification. Conversely, fire
14 may promote plant invasions via several of mechanisms (by reducing cover of competing vegetation,
15 destroying native vegetation and clearing a path for invasive plants or creating favourable soil
16 conditions) (Brooks et al., 2004; Grace et al., 2001; Keeley and Brennan, 2012).



17
18 **Figure 3.14 Difference between the number of invasive alien species (n=99, from(Bellard et al., 2013))**
19 **predicted to occur by 2050 (under A1B scenario) and current period “2000” within the dryland areas.**

20 At a regional scale, Bellard et al. (2013) predicted increasing risk in Africa and Asia, with declining
21 risk in Australia (Figure 3.14). This projection does not represent an exhaustive list of invasive alien
22 species occurring in drylands.

23 A set of four case studies in Ethiopia, Mexico, the USA and Pakistan is presented below to describe
24 the nuanced nature of invading plant species, their impact on drylands and their relationship with
25 climate change.

26 3.7.3.2. Ethiopia

27 The two invasive plants that inflict the heaviest damage to ecosystems, especially biodiversity, are the
28 annual herbaceous weed, *Parthenium hysterophorus* (*Asteraceae*) also known as Congress weed; and
29 the tree species, *Prosopis juliflora* (*Fabaceae*) also called Mesquite both originating from
30 southwestern United States to central - south America (Adkins and Shabbir, 2014). *Prosopis* was
31 introduced in the 1970s and has since spread rapidly. *Prosopis*, classified as the highest priority
32 invader in the country, is threatening livestock production and challenging the sustainability of the

1 pastoral systems. *Parthenium* is believed to have been introduced along with relief aid during the
2 debilitating droughts of the early 1980s, and a recent study reported that *the weed* has spread into 32
3 out of 34 districts in Tigray, the northernmost region of Ethiopia (Teka, 2016). A study by Etana et al.
4 (2011) indicated that *Parthenium* caused a 69% decline in the density of herbaceous species in Awash
5 National Park within a few years of introduction. In the presence of *Parthenium*, the growth and
6 development of crops is suppressed due to its allelopathic properties. McConnachie et al. (2011)
7 estimated a 28% crop loss across the country, including a 40-90% reduction in sorghum yield in
8 eastern Ethiopia alone (Tamado et al., 2002). The weed is a substantial agricultural and natural
9 resource problem and constitutes a significant health hazard (Fasil, 2011). *Parthenium* causes acute
10 allergic respiratory problems, skin dermatitis, and reportedly mutagenicity both in human and
11 livestock (Mekonnen, 2017; Patel, 2011). The eastern belt of Africa including Ethiopia presents a
12 very suitable habitat, and the weed is expected to spread further in the region in the future (Mainali et
13 al., 2015).

14 There is neither a comprehensive intervention plan nor a clear institutional mandate to deal with
15 invasive weeds, however, there are fragmented efforts involving local communities even though they
16 are clearly inadequate. The lessons learned are related to actions that have contributed to the current
17 scenario are several. First, lack of coordination and awareness - mesquite was introduced by
18 development agencies as a drought tolerant shade tree with little consideration of its invasive nature.
19 If research and development institutions had been aware, a containment strategy could have been
20 implemented early on. The second major lesson is the cost of inaction. When research and
21 development organisations did sound the alarm, but the warnings went largely unheeded, resulting in
22 the spread and buildup of two of the worst invasive plant species in the world (Fasil, 2011).

23 3.7.3.3. Mexico

24 Buffelgrass (*Cenchrus ciliaris* L.), a native species from southern Asia and East Africa, was
25 introduced into Texas and northern Mexico in the 1930s and 1940s, as it is highly productive in
26 drought conditions (Cox et al., 1988; Rao et al., 1996). In the Sonoran desert of Mexico, the
27 distribution of buffelgrass has increased exponentially, covering 1 M ha in Sonora State (Castellanos-
28 Villegas et al., 2002). Furthermore, its potential distribution extended to 53% of Sonora State and
29 12% of semiarid and arid ecosystems in Mexico (Arriaga et al., 2004). Buffelgrass has also been
30 reported as an aggressive invader in Australia and the United States resulting in altered fire cycles that
31 enhance further spread of this plant and disrupts ecosystem processes (Marshall et al., 2012; Miller et
32 al., 2010; Schlesinger et al., 2013).

33 Castellanos et al. (2016) reported that soil moisture was lower in the buffelgrass savanna cleared 35
34 years ago than in the native semi-arid shrubland, mainly during the summer. The ecohydrological
35 changes induced by buffelgrass can therefore displace native plant species over the long term.
36 Invasion by buffelgrass can also affect landscape productivity, as it is not as productive as native
37 vegetation (Franklin and Molina-Freaner, 2010). Incorporation of buffelgrass is considered a good
38 management practice by producers and the government. For this reason, no remedial actions are
39 undertaken.

40 3.7.3.4. United States

41 Sagebrush ecosystems have declined from 25 to 13 M ha since the late 1800s (Miller et al., 2011). A
42 major cause is the introduction of non-native cheatgrass (*Bromus tectorum*), which is the most prolific
43 invasive plant in the United States. Cheatgrass infests more than 10 M ha in the Great Basin and is
44 expanding every year (Balch et al., 2013). It provides a fine-textured fuel that increases the intensity,
45 frequency and spatial extent of fire (Balch et al., 2013). Historically, wildfire frequency was 60 to 110
46 years in Wyoming big sagebrush communities and has increased to five years following the
47 introduction of cheatgrass (Balch et al., 2013; Pilliod et al., 2017).

1 The conversion of the sagebrush steppe biome into to annual grassland with higher fire frequencies
2 has severely impacted livestock producers as grazing is not possible for a minimum of two years after
3 fire. Furthermore, cheatgrass and wildfires reduce critical habitat for wildlife and negatively impact
4 species richness and abundance – for example, the greater sage-grouse (*Centrocercus urophasianus*)
5 and pygmy rabbit (*Brachylagus idahoensis*) which are on the verge of being listed for federal
6 protection (Crawford et al., 2004; Larrucea and Brussard, 2008; Lockyer et al., 2015).

7 Attempts to reduce cheatgrass impacts through reseeding of both native and adapted introduced
8 species have occurred for more than 60 years (Hull and Stewart, 1949) with little success. Following
9 fire, cheatgrass becomes dominant and recovery of native shrubs and grasses is improbable,
10 particularly in relatively low elevation sites with minimal annual precipitation (less than 200 mm yr⁻¹)
11 (Davies et al., 2012; Taylor et al., 2014). Current rehabilitation efforts emphasise the use of native and
12 non-native perennial grasses, forbs, and shrubs (Bureau of Land Management, 2005). Recent
13 literature suggests that these treatments are not consistently effective at displacing cheatgrass
14 populations or re-establishing sage-grouse habitat with success varying with elevation and
15 precipitation (Arkle et al., 2014; Knutson et al., 2014). Proper post-fire grazing rest, season-of-use,
16 stocking rates, and subsequent management are essential to restore resilient sagebrush ecosystems
17 before they cross a threshold and become an annual grassland (Chambers et al., 2014; Miller et al.,
18 2011; Pellant et al., 2004). Biological soil crust protection may be an effective measure to reduce
19 cheatgrass germination, as biocrust disturbance has been shown to be a key factor promoting
20 germination of non-native grasses (Hernandez and Sandquist, 2011). Projections of increasing
21 temperature (Abatzoglou and Kolden, 2011), and observed reductions in and earlier melting of
22 snowpack in the Great Basin region (Harpold and Brooks, 2018; Mote et al., 2005) suggest that there
23 is a need to understand current and past climatic variability as this will drive wildfire and invasions of
24 annual grasses.

25 3.7.3.5. Pakistan

26 The alien plants invading local vegetation in Pakistan include *Brossentia papyrifera* (found in
27 Islamabad Capital territory), *Parthenium hysterophorus* (found in Punjab and Khyber Pakhtunkhwa
28 provinces), *Prosopis juliflora* (found all over Pakistan), *Eucalyptus camaldulensis* (found in Punjab
29 and Sindh provinces), *Salvinia* (aquatic plant widely distributed in water bodies in Sindh), *Cannabis*
30 *sativa* (found in Islamabad Capital Territory), *Lantana camara* and *Xanthium strumarium* (found in
31 upper Punjab and Khyber Pakhtunkhwa provinces) (Khan et al., 2010; Qureshi et al., 2014). Most of
32 these plants were introduced by the Forest Department decades ago for filling the gap between
33 demand and supply of timber, fuelwood and fodder. These non-native plants have some uses but their
34 disadvantages outweigh their benefits (Marwat et al., 2010; Rashid et al., 2014).

35 Besides being a source of biological pollution and a threat to biodiversity and habitat loss, the alien
36 plants reduce the land value and cause huge losses to agricultural communities (Rashid et al., 2014).
37 *Brossentia papyrifera*, commonly known as Paper Mulberry, is the root cause of inhalant pollen
38 allergy for the residents of lush green Islamabad during spring. From February to April, the pollen
39 allergy is at its peak with symptoms of severe persistent coughing with difficulty in breathing and
40 wheezing. The pollen count, although variable at different times and days, can be as high as 55,000
41 m⁻³. Early symptoms of the allergy include sneezing, itching in the eyes and skin, and blocked nose.
42 With changing climate, the onset of disease is getting earlier, and pollen count is estimated to cross
43 55,000 m⁻³ (Rashid et al., 2014). About 45% of allergic patients in the twin cities of Islamabad and
44 Rawalpindi showed positive sensitivity to the pollens (Marwat et al., 2010). Millions of rupees have
45 been spent by the Capital Development Authority on pruning and cutting of Paper Mulberry trees but
46 because of its regeneration capacity growth is regained rapidly (Rashid et al., 2014). Among other
47 invading plants, *Prosopis juliflora* has allelopathic properties, and *Eucalyptus* is known to transpire
48 huge amounts of water and deplete the soil of its nutrient elements (Qureshi et al., 2014).

1 Although Biodiversity Action Plan exists in Pakistan, it is not implemented in letter and spirit. The
 2 Quarantine Department focuses only on pests and pathogens but takes no notice of plant and animal
 3 species being imported. Also, there is no provision of checking the possible impacts of imported
 4 species on the environment (Rashid et al., 2014) and of carrying out bio assays of active allelopathic
 5 compounds of alien plants.

6 **3.7.4. Oases in Hyper-arid Areas in the Arabian Peninsula and Northern Africa**

7 Oases are isolated areas with reliable water supply from lakes and springs located in hyper-arid and
 8 arid zones (Figure 3.15). Oasis agriculture has long been the only viable crop production system
 9 throughout the hot and arid regions of the Arabian Peninsula and North Africa. Oases in hyper-arid
 10 climates are usually subject to water shortage as evapotranspiration exceeds rainfall. This often causes
 11 salinisation of soils. While many oases have persisted for several thousand years, many others have
 12 been abandoned, often in response to changes in climate or hydrologic conditions (Jones et al., 2019),
 13 providing testimony to societies' vulnerability to climatic shifts and raising concerns about similarly
 14 severe effects of anthropogenic climate change (Jones et al., 2019).



15 **Figure 3.15. Oases across the Arabian Peninsula and North Africa (alphabetically by country):**
 16 **(a) Masayrat ar Ruwajah oasis, Ad Dakhiliyah Governorate, Oman. Photo: Eike Lüdeling; (b)**
 17 **Tasselmanet oasis, Ouarzazate Province, Morocco. Photo: Abdellatif Khattabi. (c) Al-Ahsa**
 18 **oasis, Al-Ahsa Governorate, Saudi Arabia. Photo: Shijan Kaakkara; (d) Zarat oasis,**
 19 **Governorate of Gabes, Tunisia. Photo: Hamda Aloui; The use rights for (a), (b) and (d) were**
 20 **granted by copyright holders; (c) is licensed under the Creative Commons Attribution 2.0**
 21 **Generic license.**

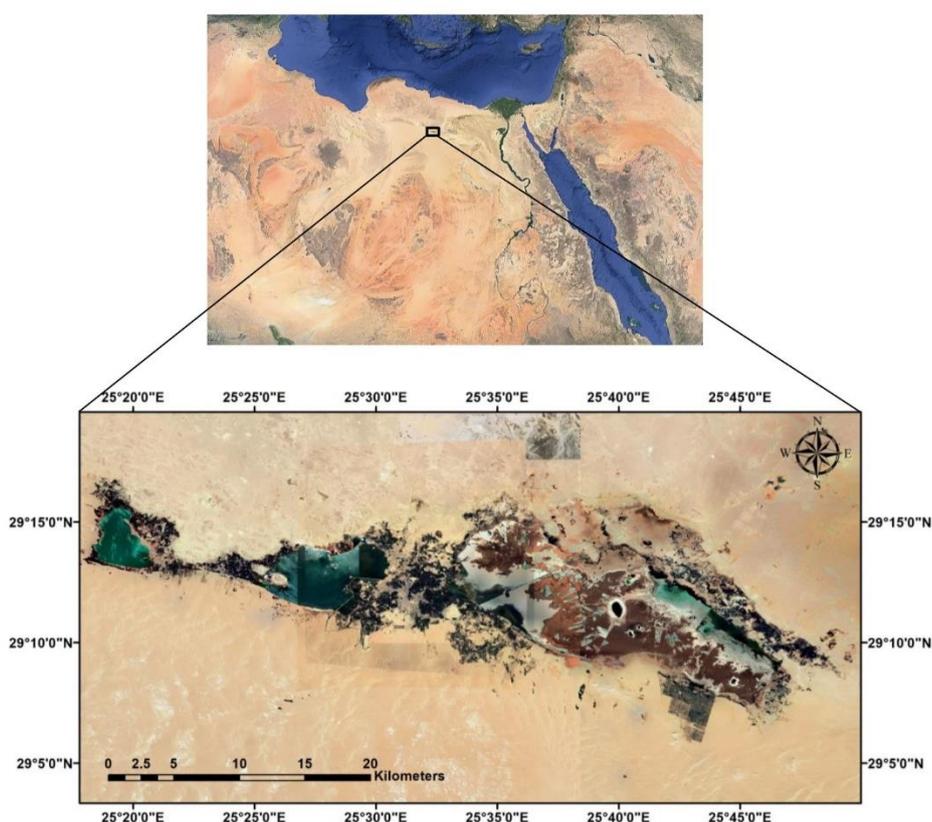
22 On the Arabian Peninsula and in North Africa, climate change is projected to have substantial and
 23 complex effects on oasis areas (Abatzoglou and Kolden, 2011; Ashkenazy et al., 2012; Bachelet et al.,

1 2016; Guan et al., 2018; Iknayan and Beissinger, 2018; Ling et al., 2013). To illustrate, by the 2050s,
2 the oases in southern Tunisia are expected to be affected by hydrological and thermal changes, with
3 an average temperature increase of 2.7°C, a 29% decrease in precipitation and a 14% increase in
4 evapotranspiration rate (Ministry of Agriculture and Water Resources of Tunisia and GIZ, 2007). In
5 Morocco, declining aquifer recharge is expected to impact the water supply of the Figuig oasis (Jilali,
6 2014), as well as for the Draa Valley (Karmaoui et al., 2016). Saudi Arabia is expected to experience
7 a 1.8–4.1°C increase in temperatures by 2050, which is forecast to raise agricultural water demand by
8 5-15% in order to maintain the level of production equal to that in 2011 (Chowdhury and Al-Zahrani,
9 2013). The increase of temperatures and variable pattern of rainfall over the central, north and south-
10 western regions of Saudi Arabia may pose challenges for sustainable water resource management
11 (Tarawneh and Chowdhury, 2018). Moreover, future climate scenarios are expected to increase the
12 frequency of floods and flash floods, such as in the coastal areas along the central parts of the Red Sea
13 and the south-southwestern areas of Saudi Arabia (Almazroui et al., 2017).

14 While many oases are cultivated with very heat-tolerant crops such as date palms, even such crops
15 eventually lose in their productivity when temperatures exceed certain thresholds or hot conditions
16 prevail for extended periods. Projections so far do not indicate severe losses in land suitability for date
17 palm for the Arabian Peninsula (Aldababseh et al., 2018; Shabani et al., 2015). It is unclear, however,
18 how reliable the climate response parameters in the underlying models are, and actual responses may
19 differ substantially. Date palms are routinely assumed to be able to endure very high temperatures, but
20 recent transcriptomic and metabolomic evidence suggests that heat stress reactions already occur at
21 35°C (Safronov et al., 2017), which is not exceptionally warm for many oases in the region. Given
22 current assumptions about the heat-tolerance of date palm, however, adverse effects are expected to
23 be small (Aldababseh et al., 2018; Shabani et al., 2015). For some other perennial oasis crops, impacts
24 of temperature increases are already apparent. Between 2004-2005 and 2012-2013, high-mountain
25 oases of Al Jabal Al Akhdar in Oman lost almost all fruit and nut trees of temperate-zone origin, with
26 the abundance of peaches, apricots, grapes, figs, pears, apples, and plums dropping by between 86%
27 and 100% (Al-Kalbani et al., 2016). This implies that that the local climate may not remain suitable
28 for species that depend on cool winters to break their dormancy period (Luedeling et al., 2009). A
29 similar impact is very probable in Tunisia and Morocco, as well as in other oasis locations in the
30 Arabian Peninsula and North Africa (Benmoussa et al., 2007). All these studies expect strong
31 decreases in winter chill, raising concerns that many currently well-established species will no longer
32 be viable in locations where they are grown today. The risk of detrimental chill shortfalls is expected
33 to increase gradually, slowly diminishing the economic prospects to produce such species. Without
34 adequate adaptation actions, the consequences of this development for many traditional oasis
35 settlements and other plantations of similar species could be highly negative.

36 At the same time, population growth and agricultural expansion in many oasis settlements are leading
37 to substantial increases in water demand for human consumption (Al-Kalbani et al., 2014). For
38 example, a large unmet water demand has been projected for future scenarios for the valley of
39 Seybouse in East Algeria (Aoun-Sebaiti et al., 2014), and similar conclusions were drawn for Wadi El
40 Natrun in Egypt (Switzman et al., 2018). Modelling studies have indicated long-term decline in
41 available water and increasing risk of water shortages, e.g. for oases in Morocco (Johannsen et al.,
42 2016; Karmaoui et al., 2016), the Dakhla oasis in Egypt's Western Desert (Sefelnasr et al., 2014) and
43 for the large Upper Mega Aquifer of the Arabian Peninsula (Siebert et al., 2016). Mainly due to the
44 risk of water shortages, Souissi et al. (2018) classified almost half of all farmers in Tunisia as non-
45 resilient to climate change, especially those relying on tree crops, which limit opportunities for short-
46 term adaptation actions.

1 The maintenance of the oasis systems and the safeguarding of their population's livelihoods are
 2 currently threatened by continuous water degradation, increasing soil salinisation, and soil
 3 contamination (Besser et al., 2017). Waterlogging and salinisation of soils due to rising saline
 4 groundwater tables coupled with inefficient drainage systems have become common to all continental
 5 oases in Tunisia, most of which are concentrated around saline depressions, known locally as chotts
 6 (Ben Hassine et al., 2013). Similar processes of salinisation are also occurring in the oasis areas of
 7 Egypt due to agricultural expansion, excessive use of water for irrigation and deficiency of the
 8 drainage systems (Abo-Ragab, 2010; Masoud and Koike, 2006). A prime example for this is Siwa
 9 oasis (Figure 3.16), a depression extending over 1050 km² in the north-western desert of Egypt in the
 10 north of the sand dune belt of the Great Sand Sea (Abo-Ragab, and Zaghoul, 2017). Siwa oasis has
 11 been recognised as a Globally Important Agricultural Heritage Site (GIAHS) by the FAO for being an
 12 *in situ* repository of plant genetic resources, especially of uniquely adapted varieties of date palm,
 13 olive and secondary crops that are highly esteemed for their quality and continue to play a significant
 14 role in rural livelihoods and diets (FAO, 2016).



15
 16 **Figure 3.16. The Satellite Image of the Siwa Oasis, Egypt. Source: Google Maps.**

17
 18 The population growth in Siwa is leading rapid agricultural expansion and land reclamation. The
 19 Siwan farmers are converting the surrounding desert into reclaimed land by applying their old
 20 inherited traditional practices. Yet, agricultural expansion in the oasis mainly depends on non-
 21 renewable groundwaters. Soil salinisation and vegetation loss have been accelerating since 2000 due
 22 to water mismanagement and improper drainage systems (Masoud and Koike, 2006). Between 1990-
 23 2008, the cultivated area increased from 53 to 88 km², lakes from 60 to 76 km², sabkhas (salt flats)
 24 from 335 to 470 km², and the urban area from 6 to 10 km² (Abo-Ragab, 2010). The problem of rising
 25 groundwater tables was exacerbated by climatic changes (Askri et al., 2010; Gad and Abdel-Baki,
 26 2002; Marlet et al., 2009).

1
2 Water supply is *likely* to become even scarcer for oasis agriculture under changing climate in the
3 future than it is today, and viable solutions are difficult to find. While some authors stress the
4 possibility to use desalinated water for irrigation (Aldababseh et al., 2018), the economics of such
5 options, especially given the high evapotranspiration rates in the Arabian Peninsula and North Africa,
6 are debatable. Many oases are located far from water sources that are suitable for desalination, adding
7 further to feasibility constraints. Most authors therefore stress the need to limit water use (Sefelnasr et
8 al., 2014), e.g. by raising irrigation efficiency (Switzman et al., 2018), reducing agricultural areas
9 (Johannsen et al., 2016) or imposing water use restrictions (Odhiambo, 2017), and to carefully
10 monitor desertification (King and Thomas, 2014). Whether adoption of crops with low water demand,
11 such as sorghum (*Sorghum bicolor* (L.) Moench) or jojoba (*Simmondsia chinensis* (Link) C. K.
12 Schneid.) (Aldababseh et al., 2018), can be a viable option for some oases remains to be seen, but
13 given their relatively low profit margins compared to currently grown oasis crops, there are reasons to
14 doubt the economic feasibility of such proposals. While it is currently unclear, to what extent oasis
15 agriculture can be maintained in hot locations of the region, cooler sites offer potential for shifting
16 towards new species and cultivars. Especially for tree crops, which have particular climatic needs
17 across seasons. Resilient options can be identified, but procedures to match tree species and cultivars
18 with site climate need to be improved to facilitate effective adaptation.

19 There is *high confidence* that many oases of North Africa and the Arabian Peninsula are vulnerable to
20 climate change. While the impacts of recent climate change are difficult to separate from the
21 consequences of other change processes, it is *likely* that water resources have already declined in
22 many places and the suitability of the local climate for many crops, especially perennial crops, has
23 already decreased. This decline of water resources and thermal suitability of oasis locations for
24 traditional crops is *very likely* to continue throughout the 21st century. In the coming years, the people
25 living in oasis regions across the world will face challenges due to increasing impacts of global
26 environmental change (Chen et al., 2018). Hence, efforts to increase their adaptive capacity to climate
27 change can facilitate the sustainable development of oasis regions globally. This will concern
28 particularly addressing the trade-offs between environmental restoration and agricultural livelihoods
29 (Chen et al., 2018). Ultimately, sustainability in oasis regions will depend on policies integrating the
30 provision of ecosystem services and social and human welfare needs (Wang et al., 2017).

31

32 **3.7.5. Integrated Watershed Management**

33 Desertification has resulted in significant loss of ecosystem processes and services as described in
34 detail in this chapter. The techniques and processes to restore degraded watersheds are not linear and
35 integrated watershed management (IWM) must address physical, biological and social approaches to
36 achieve SLM objectives (German et al., 2007).

37 **3.7.5.1. Jordan**

38 Population growth, migration into Jordan and changes in climate have resulted in desertification of the
39 Jordan Badia region. The Badia region covers more than 80% of the country's area and receives less
40 than 200 mm of rainfall per year, with some areas receiving less than 100 mm (Al-Tabini et al., 2012).
41 Climate analysis has indicated a generally increasing dryness over the West Asia and Middle Eastern
42 region (AlSarmi and Washington, 2011; Tanarhte et al., 2015) with reduction in average annual
43 rainfall in Jordan's Badia area (De Pauw et al., 2015). The incidence of extreme rainfall events has
44 not declined over the region. Locally increased incidence of extreme events over the Mediterranean
45 region have been proposed (Giannakopoulos et al., 2009).

46 The practice of intensive and localised livestock herding, in combination with deep ploughing and
47 unproductive barley agriculture, are the main drivers of severe land degradation and depletion of the
48 rangeland natural resources. This affected both the quantity and the diversity of vegetation as native

1 plants with a high nutrition value were replaced with invasive species with low palatability and
 2 nutritional content (Abu-Zanat et al., 2004). The sparsely covered and crusted soils in Jordan's Badia
 3 area have a low rainfall interception and infiltration rate, which leads to increased surface runoff and
 4 subsequent erosion and gullying, speeding up the drainage of rainwater from the watersheds that can
 5 result in downstream flooding in Amman, Jordan (Oweis, 2017).

6



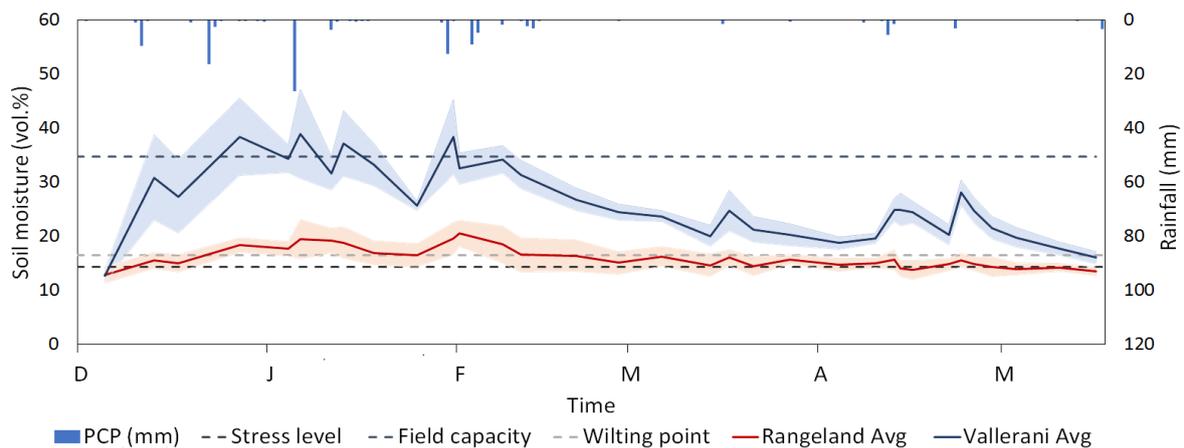
7

8 **Figure 3.17. Fresh Vallerani micro water harvesting catchment (a) and aerial imaging showing micro**
 9 **water harvesting catchment treatment after planting (b) and 1 year after treatment (c).**

10

Source: Stefan Strohmeier

11



12

13 **Figure 3.18 Illustration of enhanced soil water retention in the Mechanized Micro Rainwater Harvesting**
 14 **compared to untreated Badia rangelands in Jordan, showing precipitation (PCP), sustained stress level**
 15 **resulting in decreased production, Field Capacity and Wilting Point for available soil moisture, and then**
 16 **measured soil moisture content between the two treatments (degraded rangeland and the restored**
 17 **rangeland with the Vallerani plow).**

18 To restore the desertified Badia an IWM plan was developed using hillslope implemented water
 19 harvesting micro catchments as a targeted restoration approach (Tabieh et al., 2015). Mechanized
 20 Micro Rainwater Harvesting (MIRWH) technology using the 'Vallerani plough' (Antinori and
 21 Vallerani, 1994; Gammoh and Oweis, 2011; Ngigi, 2003) is being widely applied for rehabilitation of
 22 highly degraded rangeland areas in Jordan. Tractor digs out small water harvesting pits on the contour
 23 of the slope (Figure 3.17) allowing the retention, infiltration and the local storage of surface runoff in
 24 the soil (Oweis, 2017). The micro catchments are planted with native shrub seedlings, such as
 25 saltbush (*Atriplex halimus*), with enhanced survival as a function of increased soil moisture (Figure
 26 3.18) and increased dry matter yields (>300 kg ha⁻¹) that can serve as forage for livestock (Oweis,
 27 2017; Tabieh et al., 2015).

1 Simultaneously to MIRWH upland measures, the gully erosion is being treated through intermittent
 2 stone plug intervention (Figure 3.19), stabilising the gully beds, increasing soil moisture in proximity
 3 of the plugs and dissipating the surface runoff's energy, and mitigating further back-cutting erosion
 4 and quick drainage of water. Eventually, the treated gully areas silt up and dense vegetation cover can
 5 re-establish. In addition, grazing management practices are implemented to increase the longevity of
 6 the treatment. Ultimately, the recruitment processes and revegetation shall control the watershed's
 7 hydrological regime through rainfall interception, surface runoff deceleration and filtration, combined
 8 with the less erodible and enhanced infiltration characteristics of the rehabilitated soils. In-depth
 9 understanding of the Badia's rangeland status transition, coupled with sustainable rangeland
 10 management, are still subject to further investigation, development and adoption; required to mitigate
 11 the ongoing degradation of the Middle Eastern rangeland ecosystems.



12
 13 **Figure 3.19 Gully plug development in September 2017 (a) and post rainfall event in March 2018 (b)**
 14 **near Amman, Jordan. Source: Stefan Strohmeire.**

15 Oweis (2017) indicated that costs of the fully automated Vallerani technique was approximately USD
 16 32 ha⁻¹. The total cost of the restoration package included the production, planting, and maintenance
 17 of the shrub seedlings (USD 11 ha⁻¹). Tabieh et al. (2015) calculated a benefit cost ratio (BCR) of >
 18 1.5 for revegetation of degraded Badia areas through MIRWH and saltbush. However, costs vary
 19 based on the seedling's costs and availability of trained labour.

20 Water harvesting is not a recent scientific advancement. Water harvesting has been documented
 21 having evolved during the Bronze Age and was widely practiced in the Negev Desert during the
 22 Byzantine time period (1300-1600 years ago) (Fried et al., 2018; Stavi et al., 2017). Through
 23 construction of various structures made for packed clay and stone, water was either held on site in
 24 half-circular dam structures (Hafir) that faced up slope to capture runoff or on terraces that slowed
 25 water allowing it to infiltrate and to be stored in the soil profile. Numerous other systems were
 26 designed to capture water in below ground cisterns to be used later to provide water to livestock or for
 27 domestic use. Other water harvesting techniques divert runoff from hillslopes or wadis and spread the
 28 water in a systematic manner across playas and the toe slope of a hillslope. These systems allow
 29 production of crops in areas with 100 mm of average annual precipitation by harvesting an additional
 30 300+ mm of water (Beckers et al., 2013). Water harvesting provides a proven technology to mitigate
 31 or adapt to climate change where precipitation maybe reduced and allow for small scale crop and
 32 livestock production to continue supporting local needs.

33 **3.7.5.2. India**

34 The Green Revolution that transformed irrigated agriculture in India had little effect on agricultural
 35 productivity in the rainfed and semi-arid regions, where land degradation and drought were serious
 36 concerns. In response to this challenge, integrated watershed management (IWM) projects were
 37 implemented over large areas in semi-arid biomes over the past few decades. IWM was meant to
 38 become a key factor in meeting a range of social development goals in many semi-arid rainfed
 39 agrarian landscapes in India (Bouma et al., 2007; Kerr et al., 2002). Over the years, watershed
 40 development has become the fulcrum of rural development that has the potential to achieve the twin

1 objectives of ecosystem restoration and livelihood assurance in the drylands of India (Joy et al.,
2 2004).

3 Some reports indicate significant improvements in mitigation of drought impacts, raising crop, fodder
4 and livestock productivity, expanding the availability of drinking water and increasing incomes as a
5 result of IWM (Rao, 2000), but overall the positive impact of the programme has been questioned and
6 except in a few cases the performance has not lived up to expectations (Joy et al., 2004; JM Kerr et
7 al., 2002). Rigorous comparisons of catchments with and without IWM projects have shown no
8 significant enhancement of biomass (Bhalla et al., 2013). The factors contributing to the successful
9 cases were found to include effective participation of stakeholders in management (Rao, 2000; Ratna
10 Reddy et al., 2004) .

11 Attribution of success to soil and water conservation measures was confounded by inadequate
12 monitoring of rainfall variability and lack of catchment hydrologic indicators (Bhalla et al., 2013).
13 Social and economic trade-offs included bias of benefits to downstream crop producers at the expense
14 of pastoralists, women and upstream communities. This biased distribution of IWM benefits could
15 potentially be addressed by compensation for environmental services between communities (Kerr et
16 al., 2002). The successes in some areas also led to increased demand for water, especially
17 groundwater, since there has been no corresponding social regulation of water use after improvement
18 in water regime (Samuel et al., 2007). Policies and management did not ensure water allocation to
19 sectors with the highest social and economic benefits (Batchelor et al., 2003). Limited field evidence
20 of the positive impacts of rainwater harvesting at the local scale is available, but there are several
21 potential negative impacts at the watershed scale (Glendenning et al., 2012). Furthermore, watershed
22 projects are known to have led to more water scarcity, higher expectations for irrigation water supply,
23 further exacerbating water scarcity (Bharucha et al., 2014).

24 In summary, the overall poor performance of IWM projects have been linked to several factors. These
25 include inequity in the distribution of benefits (Kerr et al., 2002), focus on institutional aspects rather
26 than application of appropriate watershed techniques and functional aspects of watershed restoration
27 (Joy et al., 2006; Vaidyanathan, 2006), mismatch between scales of focus and those that are optimal
28 for catchment processes (Kerr, 2007), inconsistencies in criteria used to select watersheds for IWM
29 projects (Bhalla et al., 2011), and in a few cases additional costs and inefficiencies of local non-
30 governmental organisations (Chandrasekhar et al., 2006; Deshpande, 2008). Enabling policy
31 responses for improvement of IWM performance include a greater emphasis on ecological restoration
32 rather than civil engineering, sharper focus on sustainability of livelihoods than just conservation,
33 adoption of a water justice as a normative goal and minimising externalities on non-stakeholder
34 communities, rigorous independent biophysical monitoring with feedback mechanisms and
35 integration with larger schemes for food and ecological security and maintenance of environmental
36 flows for downstream areas (Bharucha et al., 2014; Calder et al., 2008; Joy et al., 2006). Successful
37 adaptation of IWM would largely depend on how IWM creatively engages with dynamics of large
38 scale land use and hydrology under a changing climate, involvement of livelihoods and rural incomes
39 in ecological restoration, regulation of groundwater use and changing aspirations of rural population
40 (*robust evidence, high agreement*) (O'Brien et al., 2004; Samuel et al., 2007; Samuel and Joy, 2018).

41 **3.7.5.3. Limpopo River Basin**

42 Covering an area of 412938 km², the Limpopo River basin spans parts of Botswana, South Africa,
43 Zimbabwe and Mozambique, eventually entering into the Mozambique Channel. It has been selected
44 as a case study as it provides a clear illustration of the combined effect of desertification and climate
45 change, and why IWM may be crucial component of reducing exposure to climate change. It is
46 predominantly a semi-arid area with an average annual rainfall of 400 mm (Mosase and Ahiablame,
47 2018). Rainfall is both highly seasonal and variable with the prominent impact of the El Nino / La
48 Nina phenomena and the Southern Oscillation leading to severe droughts (Jury, 2016). It is also
49 exposed to tropical cyclones that sweep in from the Mozambique Channel often leading to extensive
50 casualties and the destruction of infrastructure (Christie and Hanlon, 2001). Furthermore, there is
51 good agreement across climate models that the region is going to become warmer and drier, with a
52 change in the frequency of floods and droughts (Engelbrecht et al., 2011; Zhu and Ringler, 2012).

1 Seasonality is predicted to increase, which in turn may increase the frequency of flood events in an
2 area that is already susceptible to flooding (Spaliviero et al., 2014) .

3 A clear need exists to both address exposure to flood events as well as predicted decreases in water
4 availability, which are already acute. Without the additional impact of climate change, the basin is
5 rapidly reaching a point where all available water has been allocated to users (Kahinda et al., 2016;
6 Zhu and Ringler, 2012). The urgency of the situation was identified several decades ago (FAO, 2004),
7 with the countries of the Basin recognising that responses are required at several levels, both in terms
8 of system governance as well as addressing land degradation.

9 Recent reviews of the governance and implementation of IWM within the basin recognise that an
10 integrated approach is needed and that a robust institutional, legal, political, operational, technical and
11 support is crucial (Alba et al., 2016; Gbetibouo et al., 2010; Macheche et al., 2004; Spaliviero et al.,
12 2011; van der Zaag and Savenije, 1999). Within the scope of emerging lessons, two principal ones
13 emerge. The first is capacity and resource constraints at most levels. Limited capacity within
14 Limpopo Watercourse Commission (LIMCOM) and national water management authorities
15 constrains the implementation of IWM planning processes (Kahinda et al., 2016; Spaliviero et al.,
16 2011). Whereas strategy development is often relatively well-funded and resourced through donor
17 funding, long-term implementation is often limited due to competing priorities. The second is
18 adequate representation of all parties in the process in order to address existing inequalities and ensure
19 full integration of water management. For example, within Mozambique, significant strides have been
20 made towards the decentralisation of river basin governance and IWM. Despite a good progress, Alba
21 et al. (2016) found that the newly implemented system may enforce existing inequalities as not all
22 stakeholders, particularly smallholder farmers, are adequately represented in emerging water
23 management structures and are often inhibited by financial and institutional constraints. Recognising
24 economic and socio-political inequalities and explicitly considering them to ensure the representation
25 of all participants can increase the chances of successful IWM implementation.

26

27 **3.8. Knowledge Gaps and Key Uncertainties**

28 • Desertification has been studied for decades and different drivers of desertification have been
29 described, classified, and are generally understood (e.g., overgrazing by livestock or
30 salinisation from inappropriate irrigation) (D’Odorico et al., 2013). However, there are
31 knowledge gaps on the extent and severity of desertification at global, regional, and local
32 scales (Zhang and Huisingsh, 2018; Zucca et al., 2012). Overall, improved estimation and
33 mapping of areas undergoing desertification is needed. This requires a combination of rapidly
34 expanding sources of remotely sensed data, ground observations and new modelling
35 approaches. This is a critical gap, especially in the context of measuring progress towards
36 achieving the Land Degradation Neutrality target by 2030 in the framework of SDGs.

37

38 • Despite numerous relevant studies, consistent indicators for attributing desertification to
39 climatic and/or human causes are still lacking due to methodological shortcomings.

40

41 • Climate change impacts on dust and sand storm activity remain a critical gap. In addition, the
42 impacts of dust and sand storms on human welfare, ecosystems, crop productivity and animal
43 health are not measured, particularly in the highly affected regions such as the Sahel, North
44 Africa, the Middle East and Central Asia. Dust deposition on snow and ice has been found in
45 many regions of the globe (e.g. Painter et al., 2018; Kaspari et al., 2014; Qian et al., 2015;
46 Painter et al. 2013), however, the quantification of the effect globally and estimation of future
47 changes in the extent of this effect remain knowledge gaps.

48

- 1 • Future projections of combined impacts of desertification and climate change on ecosystem
2 services, fauna and flora, are lacking, even though this topic is of considerable social
3 importance. Available information is mostly on separate, individual impacts of either (mostly)
4 climate change or desertification. Responses to desertification are species-specific and
5 mechanistic models are not yet able to accurately predict individual species responses to the
6 many factors associated with desertification under changing climate.
7
- 8 • Previous studies have focused on the general characteristics of past and current desertification
9 feedbacks to the climate system, however, the information on the future interactions between
10 climate and desertification (beyond changes in the aridity index) are lacking. The knowledge
11 of future climate change impacts on such desertification processes as soil erosion, salinisation,
12 and nutrient depletion remains limited both at the global and at the local levels.
13
- 14 • Further research to develop technologies and innovations needed to combat desertification is
15 required but also better understanding of reasons for the observed poor adoption of available
16 innovations is important to improve adoption rates.
17
- 18 • Desertification under changing climate has a high potential to increase poverty particularly
19 through the risks coming from extreme weather events (Olsson et al., 2014). However, the
20 evidence rigorously attributing changes in observed poverty to climate change impacts is
21 currently not available.
22
- 23 • The knowledge on limits to adaptation to combined effects of climate change and
24 desertification is insufficient. This is an important gap since the potential for residual risks
25 and maladaptive outcomes is high.
26
- 27 • Filling these gaps involves considerable investments in research and data collection. Using
28 Earth observation systems in a standardised approach could help fill some of these gaps. This
29 would increase data comparability and reduce uncertainty in approaches and costs.
30 Systematically collected data would provide far greater insights than incomparable
31 fragmented data.
32

33 Frequently Asked Questions

34 **FAQ 3.1 How does climate change affect desertification?**

35 Desertification is land degradation in drylands. Climate change and desertification have strong
36 interactions. Desertification affects climate change through loss of fertile soil and vegetation. Soils
37 contain large amounts of carbon some of which could be released to the atmosphere due to
38 desertification, with important repercussions for the global climate system. The impacts of climate
39 change on desertification are complex and knowledge on the subject is still insufficient. On the one
40 hand, some dryland regions will receive less rainfall and increases in temperatures can reduce soil
41 moisture harming plant growth. On the other hand, the increase of CO₂ in the atmosphere can enhance
42 plant growth if there are enough water and soil nutrients available.
43

44 **FAQ 3.2 How can climate change induced desertification be avoided, reduced or reversed?**

45
46 Managing land sustainably can help avoid, reduce or reverse desertification, and contribute to climate
47 change mitigation and adaptation. Such sustainable land management practices include reducing soil
48 tillage and maintaining plant residues to keep soils covered, planting trees on degraded lands, growing
49

1 a wider variety of crops, applying efficient irrigation methods, improving rangeland grazing by
2 livestock and many others.

4 **FAQ 3.2 How do sustainable land management practices affect ecosystem services and** 5 **biodiversity?**

6 Sustainable land management practices help improve ecosystems services and protect biodiversity.
7 For example, conservation agriculture and better rangeland management can increase the production
8 of food and fibres. Planting trees on degraded lands can improve soil fertility and fix carbon in soils.
9 Sustainable land management practices also support biodiversity through habitat protection.
10 Biodiversity protection allows to safeguard precious genetic resources, thus, contributing to human
11 wellbeing.

14 **References**

- 15 Abatzoglou, J. T., Dobrowski, S. Z., Parks, S. A., & Hegewisch, K. C. (2018). TerraClimate, a high-
16 resolution global dataset of monthly climate and climatic water balance from 1958–2015.
17 *Scientific Data*, 5, 170191. <https://doi.org/10.1038/sdata.2017.191>
- 18 Abatzoglou, J. T., & Kolden, C. A. (2011). Climate Change in Western US Deserts: Potential for
19 Increased Wildfire and Invasive Annual Grasses. *Rangeland Ecology & Management*, 64(5),
20 471–478. <https://doi.org/10.2111/REM-D-09-00151.1>
- 21 Abbass, I. (2014). No retreat no surrender conflict for survival between Fulani pastoralists and
22 farmers in Northern Nigeria. *European Scientific Journal*, 8(1). Retrieved from
23 <http://www.eujournal.org/index.php/esj/article/view/4618>
- 24 Abdalla, M., Hastings, A., Chadwick, D. R. R., Jones, D. L. L., Evans, C. D. D., Jones, M. B. B., ...
25 Smith, P. (2018). Critical review of the impacts of grazing intensity on soil organic carbon
26 storage and other soil quality indicators in extensively managed grasslands. *Agriculture,*
27 *Ecosystems and Environment*, 253(May 2017), 62–81.
28 <https://doi.org/10.1016/j.agee.2017.10.023>
- 29 Abdelkebir, T., & Ferchichi, A. (2015). *Lutte biologique contre les accumulations sableuses dans la*
30 *Tunisie aride. Revue des Régions Arides.*
- 31 Abdu, N., & Robinson, L. (2017). *Community-based rangeland management in Dirre rangeland unit:*
32 *Taking Successes in Land Restoration to scale project.* Retrieved from
33 https://cgspace.cgiar.org/bitstream/handle/10568/89714/pr_dirre.pdf?sequence=6
- 34 Abdulai, A., Owusu, V., & Goetz, R. (2011). Land tenure differences and investment in land
35 improvement measures: Theoretical and empirical analyses. *Journal of Development Economics*,
36 96(1), 66–78. <https://doi.org/10.1016/J.JDEVECO.2010.08.002>
- 37 Abdullaev, I., Ul Hassan, M., & Jumaboev, K. (2007). Water saving and economic impacts of land
38 leveling: the case study of cotton production in Tajikistan. *Irrigation and Drainage Systems*,
39 21(3–4), 251–263. <https://doi.org/10.1007/s10795-007-9034-2>
- 40 Abed, R. M. M., Ramette, A., Hübner, V., de Deckker, P., & de Beer, D. (2012). Microbial diversity
41 of eolian dust sources from saline lake sediments and biological soil crusts in arid Southern
42 Australia. *FEMS Microbiology Ecology*, 80(2), 294–304. <https://doi.org/10.1111/j.1574-6941.2011.01289.x>
- 44 Abideen, Z., Hameed, A., Koyro, H., Gul, B., Ansari, R., & Khan, M. (2014). Sustainable biofuel
45 production from non-food sources? An overview. *Emirates Journal of Food and Agriculture*,
46 26(12), 1057. <https://doi.org/10.9755/ejfa.v26i12.19107>

- 1 Abo-Ragab, S., Zaghloul, A. Q. (2017). Sand dunes movement and its impact on development plans,
2 Siwa oasis, Egypt. In *Proceedings of the VIII International Agricultural Symposium. AGROSYM*
3 *2017*. .
- 4 Abo-Ragab, S. (2010). *A Desertification Impact on Siwa Oasis: Present and Future Challenges.*
5 *Research Journal of Agriculture and Biological Sciences* (Vol. 6). Retrieved from
6 <http://www.aensiweb.net/AENSIWEB/rjabs/rjabs/2010/791-805.pdf>
- 7 Aboushook, M., Maghrabi, M. N. Al, Fadol, A., & Abdelhaffez, G. S. (2012). Different methods for
8 stabilisation of sand dunes using calcium bentonite. *International Journal of Environmental*
9 *Engineering*, *4*(1/2), 79. <https://doi.org/10.1504/IJEE.2012.048103>
- 10 Abril, A., Barttfeld, P., & Bucher, E. H. (2005). The effect of fire and overgrazing disturbs on soil
11 carbon balance in the Dry Chaco forest. *Forest Ecology and Management*, *206*, 399–405.
- 12 Abu-Zanat, M. ., Ruyle, G. ., & Abdel-Hamid, N. . (2004). Increasing range production from fodder
13 shrubs in low rainfall areas. *Journal of Arid Environments*, *59*(2), 205–216.
14 <https://doi.org/10.1016/J.JARIDENV.2003.12.011>
- 15 Acevedo, E., & Silva, P. (2003). Agronomia de la cero labranza. *Serie Ciencias Agronómicas*,
16 *3*(June), 118.
- 17 Achite, M., & Ouillon, S. (2007). Suspended sediment transport in a semiarid watershed, Wadi Abd,
18 Algeria (1973–1995). *Journal of Hydrology*, *343*(3–4), 187–202.
19 <https://doi.org/10.1016/J.JHYDROL.2007.06.026>
- 20 Achite, M., Ouillon, S., Quoc Viet, H., & Giay, C. (2016). Recent changes in climate, hydrology and
21 sediment load in the Wadi Abd, Algeria (1970-2010). *Hydrol. Earth Syst. Sci*, *20*, 1355–1372.
22 <https://doi.org/10.5194/hess-20-1355-2016>
- 23 Ackerley, D., Booth, B. B. B., Knight, S. H. E., Highwood, E. J., Frame, D. J., Allen, M. R., ...
24 Rowell, D. P. (2011). Sensitivity of Twentieth-Century Sahel Rainfall to Sulfate Aerosol and
25 CO₂ Forcing. *Journal of Climate*, *24*(19), 4999–5014. [https://doi.org/10.1175/JCLI-D-11-](https://doi.org/10.1175/JCLI-D-11-00019.1)
26 [00019.1](https://doi.org/10.1175/JCLI-D-11-00019.1)
- 27 Adams, C., Ide, T., Barnett, J., & Detges, A. (2018). Sampling bias in climate–conflict research.
28 *Nature Climate Change*, *8*(3), 200–203. <https://doi.org/10.1038/s41558-018-0068-2>
- 29 Adger, W. N. (2003). Social Capital, Collective Action, and Adaptation to Climate Change. *Economic*
30 *Geography*, *79*(4), 387–404. <https://doi.org/10.1111/j.1944-8287.2003.tb00220.x>
- 31 Adger, W. N., Pulhin, J. M., Barnett, J., Dabelko, G. D., Hovelsrud, G. K., Levy, M., ... Vogel, C. H.
32 (2014). Human security. In C. B. Field, V. R. Barros, D. J. Dokken, T. E. B. K.J. Mach, M.D.
33 Mastrandrea, S. M. M. Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova, B. Girma, E.S. Kissel,
34 A.N. Levy, & and L. L. W. P.R. Mastrandrea (Eds.), *Climate Change 2014: Impacts,*
35 *Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working*
36 *Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* (pp.
37 755–791). Cambridge [England]; New York: Cambridge University Press.
- 38 Adkins, S., & Shabbir, A. (2014). Biology, ecology and management of the invasive parthenium weed
39 (*Parthenium hysterophorus* L.). *Pest Management Science*, *70*(7), 1023–1029.
40 <https://doi.org/10.1002/ps.3708>
- 41 Afsharzade, N., Papzan, A., & M Ashjaee. (2016). Renewable energy development in rural areas of
42 Iran. *Renewable and Sustainable Energy Reviews*, *65*, 743–755. Retrieved from
43 <http://www.sciencedirect.com/science/article/pii/S1364032116303720>
- 44 Agier, L., Deroubaix, A., Martiny, N., Yaka, P., Djibo, A., & Broutin, H. (2012). Seasonality of
45 meningitis in Africa and climate forcing: aerosols stand out. *Journal of The Royal Society*
46 *Interface*, *10*(79), 20120814–20120814. <https://doi.org/10.1098/rsif.2012.0814>
- 47 Aha, B., & Ayitey, J. Z. (2017). Biofuels and the hazards of land grabbing: Tenure (in)security and
48 indigenous farmers’ investment decisions in Ghana. *Land Use Policy*, *60*, 48–59.

- 1 <https://doi.org/10.1016/J.LANDUSEPOL.2016.10.012>
- 2 Ahlerup, P., Olsson, O., & Yanagizawa, D. (2009). Social capital vs institutions in the growth process.
3 *European Journal of Political Economy*, 25(1), 1–14.
4 <https://doi.org/10.1016/J.EJPOLECO.2008.09.008>
- 5 Ahmed, A., Lawson, E. T., Mensah, A., Gordon, C., & Padgham, J. (2016). Adaptation to climate
6 change or non-climatic stressors in semi-arid regions? Evidence of gender differentiation in
7 three agrarian districts of Ghana. *Environmental Development*, 20, 45–58.
8 <https://doi.org/10.1016/J.ENVDEV.2016.08.002>
- 9 Ait Hou, M., Grazia, C., & Malorgio, G. (2015). Food safety standards and international supply chain
10 organization: A case study of the Moroccan fruit and vegetable exports. *Food Control*, 55, 190–
11 199. <https://doi.org/10.1016/J.FOODCONT.2015.02.023>
- 12 Ajai, R. R., Arya, A. S., Dhinwa, A. S., Pathan, S. K., & Raj, K. G. (2009). Desertification/land
13 degradation status mapping of India. *Current Science*, 97(10), 1478–1483.
- 14 Ajayi, O. C., Akinnifesi, F. K., & Ajayi, A. O. (2016). How by-laws and collective action influence
15 farmers' adoption of agroforestry and natural resource management technologies: lessons from
16 Zambia. *Forests, Trees and Livelihoods*, 25(2), 102–113.
17 <https://doi.org/10.1080/14728028.2016.1153435>
- 18 Akay, A., & Yildirim, A. I. (2010). *The Greening Desert Of Karapinar: An Example from Turkey*.
19 Retrieved from <https://www.researchgate.net/publication/265235140>
- 20 Akça, E., Takashi, K., & Sato, T. (2016). Development and Success, For Whom and Where: The
21 Central Anatolian Case. *Land Restoration*, 533–541. <https://doi.org/10.1016/B978-0-12-801231-4.00034-3>
- 23 Akhtar-Schuster, M., Thomas, R. J., Stringer, L. C., Chasek, P., & Seely, M. (2011). Improving the
24 enabling environment to combat land degradation: Institutional, financial, legal and science-
25 policy challenges and solutions. *Land Degradation & Development*, 22(2), 299–312.
26 <https://doi.org/10.1002/ldr.1058>
- 27 Akhter, J., Mahmood, K., Malik, K. A., Ahmed, S., & Murray, R. (2003). Amelioration of a saline
28 sodic soil through cultivation of a salt-tolerant grass *Leptochloa fusca*. *Environmental*
29 *Conservation*, 30(2), 168–174. <https://doi.org/10.1017/S0376892903000158>
- 30 Akhter, R., & Arshad, M. (2006). Arid rangelands in the Cholistan Desert (Pakistan). *Article*
31 *Scientifique*, 17(1–2), 210–7. Retrieved from [http://www.jle.com/download/sec-270101-
32 arid_rangelands_in_the_cholistan_desert_pakistan_--W9nBO38AAQEAAAC319tUAAAAB-
33 a.pdf](http://www.jle.com/download/sec-270101-arid_rangelands_in_the_cholistan_desert_pakistan_--W9nBO38AAQEAAAC319tUAAAAB-a.pdf)
- 34 Akinshina, N., Azizov, A., Karasyova, T., & Klose, E. (2016). On the issue of halophytes as energy
35 plants in saline environment. *Biomass and Bioenergy*, 91, 306–311.
36 <https://doi.org/10.1016/J.BIOMBIOE.2016.05.034>
- 37 Al-Bakri, J. T., Brown, L., Gedalof, Z., Berg, A., Nickling, W., Khresat, S., ... Saoub, H. (2016).
38 Modelling desertification f in the north-west of Jordan using geospatial and remote sensing
39 techniques. *Geomatics, Natural Hazards and Risk*, 7(2), 531–549.
40 <https://doi.org/10.1080/19475705.2014.945102>
- 41 Al-Kalbani, M., Price, M., Abahussain, A., Ahmed, M., O'Higgins, T., Al-Kalbani, M. S., ...
42 O'Higgins, T. (2014). Vulnerability Assessment of Environmental and Climate Change Impacts
43 on Water Resources in Al Jabal Al Akhdar, Sultanate of Oman. *Water*, 6(10), 3118–3135.
44 <https://doi.org/10.3390/w6103118>
- 45 Al-Kalbani, M. S., Price, M. F., O'Higgins, T., Ahmed, M., & Abahussain, A. (2016). Integrated
46 environmental assessment to explore water resources management in Al Jabal Al Akhdar,
47 Sultanate of Oman. *Regional Environmental Change*, 16(5), 1345–1361.
48 <https://doi.org/10.1007/s10113-015-0864-4>

- 1 Al-Mulla, L., Bhat, N. R., & Khalil, M. (2013). Salt-Tolerance of Tissue-Cultured Date Palm
2 Cultivars under Controlled Environment. *World Academy of Science, Engineering and*
3 *Technology International Journal of Animal and Veterinary Sciences*, 7(8), 811–814. Retrieved
4 from [https://waset.org/publications/16217/salt-tolerance-of-tissue-cultured-date-palm-cultivars-](https://waset.org/publications/16217/salt-tolerance-of-tissue-cultured-date-palm-cultivars-under-controlled-environment)
5 [under-controlled-environment](https://waset.org/publications/16217/salt-tolerance-of-tissue-cultured-date-palm-cultivars-under-controlled-environment)
- 6 Al-Tabini, R., Al-Khalidi, K., & Al-Shudiefat, M. (2012). Livestock, medicinal plants and rangeland
7 viability in Jordan's Badia: through the lens of traditional and local knowledge. *Pastoralism:*
8 *Research, Policy and Practice*, 2(1), 4. <https://doi.org/10.1186/2041-7136-2-4>
- 9 Al Hamndou, D., & Requier-Desjardins, M. (2008). Variabilité climatique, désertification et
10 biodiversité en afrique : s'adapter, une approche intégrée. *Vertigo*, (Volume 8 Numéro 1).
11 <https://doi.org/10.4000/vertigo.5356>
- 12 Aladejana, O. O., Salami, A. T., & Adetoro, O.-I. O. (2018). Hydrological responses to land
13 degradation in the Northwest Benin Owena River Basin, Nigeria. *Journal of Environmental*
14 *Management*, 225, 300–312. <https://doi.org/10.1016/J.JENVMAN.2018.07.095>
- 15 Alam, K. (2015). Farmers' adaptation to water scarcity in drought-prone environments: A case study
16 of Rajshahi District, Bangladesh. *Agricultural Water Management*, 148, 196–206.
17 <https://doi.org/10.1016/J.AGWAT.2014.10.011>
- 18 Alba, R., Bolding, A., & Ducrot, R. (2016). The Politics of Water Payments and Stakeholder
19 Participation in the Limpopo River Basin, Mozambique Raphaëlle Ducrot. *Water Alternatives*,
20 9(3), 569–587. Retrieved from www.water-alternatives.org
- 21 Albaladejo, J., Ortiz, R., Garcia-Franco, N., Navarro, A. R., Almagro, M., Pintado, J. G., & Martínez-
22 Mena, M. (2013). Land use and climate change impacts on soil organic carbon stocks in semi-
23 arid Spain. *Journal of Soils and Sediments*, 13(2), 265–277. <https://doi.org/10.1007/s11368-012->
24 [0617-7](https://doi.org/10.1007/s11368-012-0617-7)
- 25 Albanito, F., Beringer, T., Corstanje, R., Poulter, B., Stephenson, A., Zawadzka, J., & Smith, P.
26 (2016). Carbon implications of converting cropland to bioenergy crops or forest for climate
27 mitigation: a global assessment. *GCB Bioenergy*, 8(1), 81–95.
28 <https://doi.org/10.1111/gcbb.12242>
- 29 Albright, T. P., Mutiibwa, D., Gerson, A. R., Smith, E. K., Talbot, W. A., O'Neill, J. J., ... & Wolf, B.
30 O. (2017). Mapping evaporative water loss in desert passerines reveals an expanding threat of
31 lethal dehydration. *Proceedings of the National Academy of Sciences*, 114(9), 2283–2288.
32 <https://doi.org/10.1073/pnas.1613625114>
- 33 Aldababseh, A., Temimi, M., Maghelal, P., Branch, O., Wulfmeyer, V., Aldababseh, A., ...
34 Wulfmeyer, V. (2018). Multi-Criteria Evaluation of Irrigated Agriculture Suitability to Achieve
35 Food Security in an Arid Environment. *Sustainability*, 10(3), 803.
36 <https://doi.org/10.3390/su10030803>
- 37 Alessandri, A., De Felice, M., Zeng, N., Mariotti, A., Pan, Y., Cherchi, A., ... Artale, V. (2014).
38 Robust assessment of the expansion and retreat of Mediterranean climate in the 21st century.
39 *Scientific Reports*, 4, 4–11. <https://doi.org/10.1038/srep07211>
- 40 Alimaev, I. I., Kerven, C., Torekhanov, A., Behnke, R., Smailov, K., Yurchenko, V., ... Shanbaev, K.
41 (2008). The Impact of Livestock Grazing on Soils and Vegetation Around Settlements in
42 Southeast Kazakhstan. In *The Socio-Economic Causes and Consequences of Desertification in*
43 *Central Asia* (pp. 81–112). Dordrecht: Springer Netherlands. <https://doi.org/10.1007/978-1->
44 [4020-8544-4_5](https://doi.org/10.1007/978-1-4020-8544-4_5)
- 45 Aliyu, A., Modu, B., & C.W. Tan. (2017). A review of renewable energy development in Africa: A
46 focus in South Africa, Egypt and Nigeria. *Renewable and Sustainable Energy Reviews*, 81(2).
47 Retrieved from <https://www.sciencedirect.com/science/article/pii/S1364032117309887>
- 48 Alkire, S., & Santos, M. E. (2010). *Acute multidimensional poverty: a new index for developing*
49 *countries*. University of Oxford, Poverty and Human Development Initiative. Retrieved from

- 1 <http://ophi.org.uk/acute-multidimensional-poverty-a-new-index-for-developing-countries/>
- 2 Alkire, S., & Santos, M. E. (2014). Measuring Acute Poverty in the Developing World: Robustness
3 and Scope of the Multidimensional Poverty Index. *World Development*, 59, 251–274.
4 <https://doi.org/10.1016/J.WORLDDEV.2014.01.026>
- 5 Almagro, M., López, J., Boix-Fayos, C., Albaladejo, J., & Martínez-Mena, M. (2010). Belowground
6 carbon allocation patterns in a dry Mediterranean ecosystem: A comparison of two models. *Soil*
7 *Biology and Biochemistry*, 42(9), 1549–1557. <https://doi.org/10.1016/j.soilbio.2010.05.031>
- 8 Almazroui, M., Alobaidi, M., Saeed, S., Mashat, A., & Assiri, M. (2018). The possible impact of the
9 circumglobal wave train on the wet season dust storm activity over the northern Arabian
10 Peninsula. *Climate Dynamics*, 50(5–6), 2257–2268. <https://doi.org/10.1007/s00382-017-3747-1>
- 11 Almazroui, M., Islam, M. N., Balkhair, K. S., Şen, Z., & Masood, A. (2017). Rainwater harvesting
12 possibility under climate change: A basin-scale case study over western province of Saudi
13 Arabia. *Atmospheric Research*, 189, 11–23. <https://doi.org/10.1016/J.ATMOSRES.2017.01.004>
- 14 Almer, C., Laurent-Lucchetti, J., & Oechlin, M. (2017). Water scarcity and rioting: Disaggregated
15 evidence from Sub-Saharan Africa. *Journal of Environmental Economics and Management*, 86,
16 193–209. <https://doi.org/10.1016/J.JEEM.2017.06.002>
- 17 Alobaidi, M., Almazroui, M., Mashat, A., & Jones, P. D. (2017). Arabian Peninsula wet season dust
18 storm distribution: regionalization and trends analysis (1983-2013). *International Journal of*
19 *Climatology*, 37(3), 1356–1373. <https://doi.org/10.1002/joc.4782>
- 20 Alrasbi, S. A. R., Hussain, N., & Schmeisky, H. (2010). Evaluation of the growth of date palm
21 seedlings irrigated with saline water in the Sultanate of Oman. In *IV International Date Palm*
22 *Conference* (pp. 233–246).
- 23 AlSarmi, S., & Washington, R. (2011). Recent observed climate change over the Arabian Peninsula.
24 *Journal of Geophysical Research*, 116(D11), D11109. <https://doi.org/10.1029/2010JD015459>
- 25 Altchenko, Y., & Villholth, K. G. (2015). Mapping irrigation potential from renewable groundwater
26 in Africa – a quantitative hydrological approach. *Hydrology and Earth System Sciences*, 19(2),
27 1055–1067. <https://doi.org/10.5194/hess-19-1055-2015>
- 28 Altieri, M. A. (2018). *Agroecology : the science of sustainable agriculture*. Boca Raton: CRC Press.
- 29 Altieri, M. A., & Nicholls, C. I. (2017). The adaptation and mitigation potential of traditional
30 agriculture in a changing climate. *Climatic Change*, 140(1), 33–45.
31 <https://doi.org/10.1007/s10584-013-0909-y>
- 32 Altieri, M. A., Nicholls, C. I., Henao, A., & Lana, M. A. (2015). Agroecology and the design of
33 climate change-resilient farming systems. *Agronomy for Sustainable Development*, 35(3), 869–
34 890. <https://doi.org/10.1007/s13593-015-0285-2>
- 35 Altieri, M., & Koohafkan, P. (2008). *Enduring farms: climate change, smallholders and traditional*
36 *farming communities*. Penang, Malaysia: Third World Network. Retrieved from
37 http://sa.indiaenvironmentportal.org.in/files/Enduring_Farms.pdf
- 38 Amadou, A., Abdourhamane Touré, A., Rajot, J., Garba, Z., Marticorena, B., Petit, C., & Sebag, D.
39 (2011). *Impact of very low crop residues cover on wind erosion in the Sahel*. *Catena* (Vol. 85).
40 <https://doi.org/10.1016/j.catena.2011.01.002>
- 41 Amiraslani, F., & Dragovich, D. (2011). Combating desertification in Iran over the last 50 years: An
42 overview of changing approaches. *Journal of Environmental Management*, 92(1), 1–13.
43 <https://doi.org/10.1016/j.jenvman.2010.08.012>
- 44 Amugune, I., Cerutti, P. O., Baral, H., Leonard, S., & Martius, C. (2017). *Small flame but no fire:*
45 *Wood fuel in the (Intended) Nationally Determined Contributions of countries in Sub-Saharan*
46 *Africa* (CIFOR Working Paper No. 232). Center for International Forestry Research (CIFOR),
47 Bogor, Indonesia. <https://doi.org/10.17528/CIFOR/006651>

- 1 Amundson, R., Berhe, A., Hopmans, J., Olson, C., Ester Szein, A., & L Sparks, D. (2015). *Soil*
2 *science. Soil and human security in the 21st century. Science* (Vol. 348).
3 <https://doi.org/10.1126/science.1261071>
- 4 An, P., Inanaga, S., Zhu, N., Li, X., Fadul, H. M., & Mars, M. (2007). Plant species as indicators of
5 the extent of desertification in four sandy rangelands. *African Journal of Ecology*, 45(1), 94–
6 102. <https://doi.org/10.1111/j.1365-2028.2006.00681.x>
- 7 Anache, J. A. A., Flanagan, D. C., Srivastava, A., & Wendland, E. C. (2018). Land use and climate
8 change impacts on runoff and soil erosion at the hillslope scale in the Brazilian Cerrado. *Science*
9 *of The Total Environment*, 622–623, 140–151.
10 <https://doi.org/10.1016/J.SCITOTENV.2017.11.257>
- 11 Andela, N., Liu, Y. Y., van Dijk, A. I. J. M., de Jeu, R. A. M., & McVicar, T. R. (2013). Global
12 changes in dryland vegetation dynamics (1988–2008) assessed by satellite remote sensing:
13 comparing a new passive microwave vegetation density record with reflective greenness data.
14 *Biogeosciences*, 10(10), 6657–6676. <https://doi.org/10.5194/bg-10-6657-2013>
- 15 Andersson, K. P., & Ostrom, E. (2008). Analyzing decentralized resource regimes from a polycentric
16 perspective. *Policy Sciences*, 41(1), 71–93. <https://doi.org/10.1007/s11077-007-9055-6>
- 17 Angel, S., Parent, J., Civco, D., & Blei, A. (2011). *Making room for a planet of cities*. Retrieved from
18 [https://www.lincolnst.edu/sites/default/files/pubfiles/making-room-for-a-planet-of-cities-](https://www.lincolnst.edu/sites/default/files/pubfiles/making-room-for-a-planet-of-cities-full_0.pdf)
19 [full_0.pdf](https://www.lincolnst.edu/sites/default/files/pubfiles/making-room-for-a-planet-of-cities-full_0.pdf)
- 20 Antinori, P., & Vallerani, V. (1994). Experiments in water harvesting technology with the dolphin and
21 train ploughs. In FAO (Ed.), *Water harvesting for improved agricultural production* (pp. 113–
22 132). Rome.
- 23 Antwi-Agyei, P., Dougill, A. J., Stringer, L. C., & Codjoe, S. N. A. (2018). Adaptation opportunities
24 and maladaptive outcomes in climate vulnerability hotspots of northern Ghana. *Climate Risk*
25 *Management*, 19, 83–93. <https://doi.org/10.1016/J.CRM.2017.11.003>
- 26 Anyamba, A., & Tucker, C. J. (2005). Analysis of Sahelian vegetation dynamics using NOAA-
27 AVHRR NDVI data from 1981–2003. *Journal of Arid Environments*, 63(3), 596–614.
28 <https://doi.org/10.1016/J.JARIDENV.2005.03.007>
- 29 Aoun-Sebaiti, B., Hani, A., Djabri, L., Chaffai, H., Aichouri, I., & Boughrira, N. (2014). Simulation
30 of water supply and water demand in the valley of Seybouse (East Algeria). *Desalination and*
31 *Water Treatment*, 52(10–12), 2114–2119. <https://doi.org/10.1080/19443994.2013.855662>
- 32 Apuri, I., Peprah, K., & Achana, G. T. W. (2018). Climate change adaptation through agroforestry:
33 The case of Kassena Nankana West District, Ghana. *Environmental Development*, 28, 32–41.
34 <https://doi.org/10.1016/J.ENVDEV.2018.09.002>
- 35 Archer, S. R., Andersen, E. M., Predick, K. I., Schwinning, S., Steidl, R. J., & Woods, S. R. (2017).
36 Woody plant encroachment: causes and consequences. In D. D. Briske (Ed.), *Rangeland*
37 *Systems*. Springer Series on Environmental Management. [https://doi.org/10.1007/978-3-319-](https://doi.org/10.1007/978-3-319-46709-2)
38 [46709-2](https://doi.org/10.1007/978-3-319-46709-2)
- 39 Archibald, S., Roy, D. P., Van Wilgen, B. W., & Scholes, R. J. (2009). What limits fire? An
40 examination of drivers of burnt area in southern Africa. *Global Change Biology*, 15, 613–630.
41 Retrieved from <http://dx.doi.org/10.1111/j.1365-2486.2008.01754.x>
- 42 Arkle, R. S., Pilliod, D. S., Hanser, S. E., Brooks, M. L., Chambers, J. C., Grace, J. B., ... Wirth, T.
43 A. (2014). Quantifying restoration effectiveness using multi-scale habitat models: implications
44 for sage-grouse in the Great Basin. *Ecosphere*, 5(3), art31. <https://doi.org/10.1890/ES13-00278.1>
- 45 Arora-Jonsson, S. (2011). Virtue and vulnerability: Discourses on women, gender and climate change.
46 *Global Environmental Change*, 21(2), 744–751.
47 <https://doi.org/10.1016/J.GLOENVCHA.2011.01.005>
- 48 Arriaga, L., Castellanos V., A. E., Moreno, E., & Alarcón, J. (2004). Potential Ecological Distribution

- 1 of Alien Invasive Species and Risk Assessment: a Case Study of Buffel Grass in Arid Regions
2 of Mexico. *Conservation Biology*, 18(6), 1504–1514. <https://doi.org/10.1111/j.1523->
3 1739.2004.00166.x
- 4 Asadi Zarch, M. A., Sivakumar, B., & Sharma, A. (2015). Assessment of global aridity change.
5 *Journal of Hydrology*, 520, 300–313. <https://doi.org/10.1016/J.JHYDROL.2014.11.033>
- 6 Asem, S. O., & Roy, W. Y. (2010). Biodiversity and climate change in Kuwait.
7 <https://doi.org/10.1108/17568691011020265>
- 8 Ashkenazy, Y., Yizhaq, H., & Tsoar, H. (2012). Sand dune mobility under climate change in the
9 Kalahari and Australian deserts. *Climatic Change*, 112(3–4), 901–923.
10 <https://doi.org/10.1007/s10584-011-0264-9>
- 11 Ashraf, M. Y., Ashraf, M., Mahmood, K., Akhter, J., Hussain, F., & Arshad, M. (2010).
12 Phytoremediation of Saline Soils for Sustainable Agricultural Productivity. In *Plant Adaptation*
13 *and Phytoremediation* (pp. 335–355). Dordrecht: Springer Netherlands.
14 https://doi.org/10.1007/978-90-481-9370-7_15
- 15 Askri, B., Bouhlila, R., & Job, J. O. (2010). Development and application of a conceptual hydrologic
16 model to predict soil salinity within modern Tunisian oases. *Journal of Hydrology*, 380(1–2),
17 45–61. <https://doi.org/10.1016/J.JHYDROL.2009.10.022>
- 18 Aslam, M., Prathapar, S. A. (2006). *Strategies to Mitigate Secondary Salinization in the Indus Basin*
19 *of Pakistan: A selective review*. Retrieved from <https://core.ac.uk/download/pdf/6404971.pdf>
- 20 Asner, G. P., Archer, S., Hughes, R. F., Ansley, R. J., & Wessman, C. A. (2003). Net changes in
21 regional woody vegetation cover and carbon storage in Texas Drylands, 1937-1999. *Global*
22 *Change Biology*, 9, 316–335.
- 23 Asner, G. P., Knapp, D. E., Boardman, J., Green, R. O., Kennedy-Bowdoin, T., Eastwood, M., ...
24 Field, C. B. (2012). Carnegie Airborne Observatory-2: Increasing science data dimensionality
25 via high-fidelity multi-sensor fusion. *Remote Sensing of Environment*, 124, 454–465.
26 <https://doi.org/10.1016/J.RSE.2012.06.012>
- 27 Auclair, L., Baudot, P., Genin, D., Romagny, B., & Simenel, R. (2011). Patrimony for Resilience
28 Evidence from the Forest Agdal in the Moroccan High Atlas. *Ecology and Society*, 16(4).
29 <https://doi.org/10.5751/ES-04429-160424>
- 30 Austin, A. T., Yahdjian, L., Stark, J. M., Belnap, J., Porporato, A., Norton, U., ... Schaeffer, S. M.
31 (2004). Water pulses and biogeochemical cycles in arid and semiarid ecosystems. *Oecologia*,
32 141(2), 221–235. <https://doi.org/10.1007/s00442-004-1519-1>
- 33 Aw-Hassan, A., Korol, V., Nishanov, N., Djanibekov, U., Dubovyk, O., & Mirzabaev, A. (2016).
34 Economics of Land Degradation and Improvement in Uzbekistan. In *Economics of Land*
35 *Degradation and Improvement – A Global Assessment for Sustainable Development* (pp. 651–
36 682). Cham: Springer International Publishing. https://doi.org/10.1007/978-3-319-19168-3_21
- 37 Awan, A. R., & Mahmood, K. (2017). Tree plantation in problem soils. In M. Tahir Siddiqui and F.
38 Nawaz (Ed.), *Text Book of Applied Forestry* (pp. 140–159). Faisalabad, Pakistan: University of
39 Agriculture, Faisalabad, Pakistan.
- 40 Awotide, B. A., Karimov, A. A., & Diagne, A. (2016). Agricultural technology adoption,
41 commercialization and smallholder rice farmers' welfare in rural Nigeria. *Agricultural and Food*
42 *Economics*, 4(1), 3. <https://doi.org/10.1186/s40100-016-0047-8>
- 43 Ayanu, Y., Jentsch, A., Müller-Mahn, D., Rettberg, S., Romankiewicz, C., & Koellner, T. (2015).
44 Ecosystem engineer unleashed: *Prosopis juliflora* threatening ecosystem services? *Regional*
45 *Environmental Change*, 15(1), 155–167. <https://doi.org/10.1007/s10113-014-0616-x>
- 46 Aynekulu, E., Lohbeck, M., Nijbroek, R. P., Ordóñez, J., Turner, K. G., Vågen, T.-G., & Winowiecki,
47 L. (2017). *Review of methodologies for land degradation neutrality baselines Sub-national case*
48 *studies from Costa Rica and Namibia*. Retrieved from <http://hdl.handle.net/10568/80563>

- 1 Ayoub, A. T. (1998). Extent, severity and causative factors of land degradation in the Sudan. *Journal*
2 *of Arid Environments*, 38(3), 397–409. <https://doi.org/10.1006/JARE.1997.0346>
- 3 Bachelet, D., Ferschweiler, K., Sheehan, T., & Strittholt, J. (2016). Climate change effects on
4 southern California deserts. *Journal of Arid Environments*, 127, 17–29.
5 <https://doi.org/10.1016/J.JARIDENV.2015.10.003>
- 6 Bai, Z. G., Dent, D. L., Olsson, L., & Schaepman, M. E. (2008). Proxy global assessment of land
7 degradation. *Soil Use and Management*, 24(3), 223–234. <https://doi.org/10.1111/j.1475->
8 [2743.2008.00169.x](https://doi.org/10.1111/j.1475-2743.2008.00169.x)
- 9 Baker, L., Newell, P., & Phillips, J. (2014). The Political Economy of Energy Transitions: The Case
10 of South Africa. *New Political Economy*, 19(6), 791–818.
11 <https://doi.org/10.1080/13563467.2013.849674>
- 12 Balch, J. K., Bradley, B. A., D'Antonio, C. M., & Gómez-Dans, J. (2013). Introduced annual grass
13 increases regional fire activity across the arid western USA (1980–2009). *Global Change*
14 *Biology*, 19(1), 173–183. <https://doi.org/10.1111/gcb.12046>
- 15 Bambio, Y., & Bouayad Agha, S. (2018). Land tenure security and investment: Does strength of land
16 right really matter in rural Burkina Faso? *World Development*, 111, 130–147.
17 <https://doi.org/10.1016/J.WORLDDEV.2018.06.026>
- 18 Banadda, N. (2010). Gaps, barriers and bottlenecks to sustainable land management (SLM) adoption
19 in Uganda. *African Journal of Agricultural Research*, 5(25), 3571–3580.
20 <https://doi.org/10.5897/ajar10.029>
- 21 Bang, G. (2010). Energy security and climate change concerns: Triggers for energy policy change in
22 the United States? *Energy Policy*, 38(4), 1645–1653.
23 <https://doi.org/10.1016/J.ENPOL.2009.01.045>
- 24 Barbier, B., Yacouba, H., Karambiri, H., Zoromé, M., & Somé, B. (2009). Human Vulnerability to
25 Climate Variability in the Sahel: Farmers' Adaptation Strategies in Northern Burkina Faso.
26 *Environmental Management*, 43(5), 790–803. <https://doi.org/10.1007/s00267-008-9237-9>
- 27 Barbier, E. B., & Hochard, J. P. (2016). Does Land Degradation Increase Poverty in Developing
28 Countries? *PLOS ONE*, 11(5), e0152973. <https://doi.org/10.1371/journal.pone.0152973>
- 29 Barbier, E. B., & Hochard, J. P. (2018). Land degradation and poverty. *Nature Sustainability*, 1(11),
30 623–631. <https://doi.org/10.1038/s41893-018-0155-4>
- 31 Barger, N. N., Archer, S. R., Campbell, J. L., Huang, C. Y., Morton, J. A., & Knapp, A. K. (2011).
32 Woody plant proliferation in North American drylands: A synthesis of impacts on ecosystem
33 carbon balance. *Journal of Geophysical Research: Biogeosciences*, 116(3).
34 <https://doi.org/10.1029/2010JG001506>
- 35 Barnett, J., Evans, L. S., Gross, C., Kiem, A. S., Kingsford, R. T., Palutikof, J. P., ... Smithers, S. G.
36 (2015). From barriers to limits to climate change adaptation: path dependency and the speed of
37 change. <https://doi.org/10.5751/ES-07698-200305>
- 38 Barrett, C. B., Christiaensen, L., Sheahan, M., & Shimeles, A. (2017). On the Structural
39 Transformation of Rural Africa. *Journal of African Economies*, 26(suppl_1), i11–i35.
40 <https://doi.org/10.1093/jae/ejx009>
- 41 Barrios, S., Bertinelli, L., & Strobl, E. (2006). Climatic change and rural–urban migration: The case
42 of sub-Saharan Africa. *Journal of Urban Economics*, 60(3), 357–371.
43 <https://doi.org/10.1016/J.JUE.2006.04.005>
- 44 Baskan, O., Dengiz, O., & Demirag, İ. T. (2017). The land productivity dynamics trend as a tool for
45 land degradation assessment in a dryland ecosystem. *Environmental Monitoring and*
46 *Assessment*, 189(5), 212. <https://doi.org/10.1007/s10661-017-5909-3>
- 47 Basupi, L., Quinn, C., & AJ Dougill. (2017). Pastoralism and Land Tenure Transformation in Sub-

- 1 Saharan Africa: Conflicting Policies and Priorities in Ngamiland, Botswana. *Land*, 6(4).
2 Retrieved from <http://www.mdpi.com/2073-445X/6/4/89>
- 3 Batchelor, C. H., Rama Mohan Rao, M. S., & Manohar Rao, S. (2003). Watershed development: A
4 solution to water shortages in semi-arid India or part of the problem? *Land Use and Water*
5 *Resources Research*, 3. Retrieved from <https://ageconsearch.umn.edu/record/47866/>
- 6 Baumgartner, P. (2017). *The impacts of large-scale land-acquisition in East Africa on poverty*
7 *reduction and the rural economy: Studies in Ethiopia and Uganda*. University of Bonn.
8 Retrieved from <http://hss.ulb.uni-bonn.de/2017/4661/4661a.pdf>
- 9 Bazza, M., Kay, M., & Knutson, C. (2018). *Drought characteristics and management in North Africa*
10 *and the Near East*.
- 11 Becerril-Pina Rocio, Mastachi-Loza, C. A., Gonzalez-Sosa, E., Diaz-Delgado, C., Ba, K. M., Becerril-
12 Piña, R., ... Ba, K. M. (2015). Assessing desertification risk in the semi-arid highlands of central
13 Mexico. *Journal of Arid Environments*, 120, 4–13.
14 <https://doi.org/10.1016/j.jaridenv.2015.04.006>
- 15 Beckers, B., Berking, J., & Schütt, B. (2013). Ancient Water Harvesting Methods in the Drylands of
16 the Mediterranean and Western Asia. *ETopoi. Journal for Ancient Studies*, 2.
- 17 Beegum, S. N., Gherboudj, I., Chaouch, N., Couvidat, F., Menut, L., & Ghedira, H. (2016).
18 Simulating aerosols over Arabian Peninsula with CHIMERE: Sensitivity to soil, surface
19 parameters and anthropogenic emission inventories. *Atmospheric Environment*, 128, 185–197.
20 <https://doi.org/10.1016/J.ATMOSENV.2016.01.010>
- 21 Beegum, S. N., Gherboudj, I., Chaouch, N., Temimi, M., & Ghedira, H. (2018). Simulation and
22 analysis of synoptic scale dust storms over the Arabian Peninsula. *Atmospheric Research*, 199,
23 62–81. <https://doi.org/10.1016/J.ATMOSRES.2017.09.003>
- 24 Behnke, R. (1994). Natural Resource Management in Pastoral Africa. *Development Policy Review*,
25 12(1), 5–28. <https://doi.org/10.1111/j.1467-7679.1994.tb00053.x>
- 26 Behnke, R., & Kerven, C. (2013). Counting the costs: replacing pastoralism with irrigated agriculture
27 in the Awash Valley, north-eastern Ethiopia. In A. Catley, J. Lind, & I. Scoones (Eds.),
28 *Pastoralism and Development in Africa Dynamic Change at the Margins*. Routledge.
- 29 Behnke, R., & Mortimore, M. (2016). Introduction: The End of Desertification? In R. Behnke & M.
30 Mortimore (Eds.), *The End of Desertification? Disputing Environmental Change in the*
31 *Drylands* (pp. 1–34). Springer, Berlin, Heidelberg. https://doi.org/10.1007/978-3-642-16014-1_1
- 32 Bekchanov, M., Lamers, J. P. A., Martius, C., Bekchanov, M., Lamers, J. P. A., & Martius, C. (2010).
33 Pros and Cons of Adopting Water-Wise Approaches in the Lower Reaches of the Amu Darya: A
34 Socio-Economic View. *Water*, 2(2), 200–216. <https://doi.org/10.3390/w2020200>
- 35 Bekchanov, M., Ringler, C., Bhaduri, A., & Jeuland, M. (2016). Optimizing irrigation efficiency
36 improvements in the Aral Sea Basin. *Water Resources and Economics*, 13, 30–45.
37 <https://doi.org/10.1016/J.WRE.2015.08.003>
- 38 Belaaz, M. (2003). Le barrage vert en tant que patrimoine naturel national et moyen de lutte contre la
39 désertification. Retrieved May 25, 2018, from <http://www.fao.org/docrep/article/wfc/xii/0301->
40 [b3.htm](http://www.fao.org/docrep/article/wfc/xii/0301-b3.htm)
- 41 Belala, F., Hirche, A., Muller, S. D., Tourki, M., Salamani, M., Grandi, M., ... Boughani, M. (2018).
42 Rainfall patterns of Algerian steppes and the impacts on natural vegetation in the 20th century.
43 *Journal of Arid Land*, 10(4), 561–573. <https://doi.org/10.1007/s40333-018-0095-x>
- 44 Belay, A., Recha, J. W., Woldeamanuel, T., & Morton, J. F. (2017). Smallholder farmers' adaptation
45 to climate change and determinants of their adaptation decisions in the Central Rift Valley of
46 Ethiopia. *Agriculture & Food Security*, 6(1), 24. <https://doi.org/10.1186/s40066-017-0100-1>
- 47 Belfer, E., Ford, J. D., & Maillet, M. (2017). Representation of Indigenous peoples in climate change

- 1 reporting. *Climatic Change*, 145(1–2), 57–70. <https://doi.org/10.1007/s10584-017-2076-z>
- 2 Bellard, C., Thuiller, W., Leroy, B., Genovesi, P., Bakkenes, M., & Courchamp, F. (2013). Will
3 climate change promote future invasions? *Global Change Biology*, 19(12), 3740–3748.
4 <https://doi.org/10.1111/gcb.12344>
- 5 Ben Hassine, H., Ben Slimane, A., Mlawah, M., Albouchi, L., & Gandouzi, A. (2013). Effects of
6 Underground Water on Soil Salinity and Dates Production in Kebili Oases Area (Tunisia): The
7 Case of El Bahaier Oasis. *IOSR Journal of Environmental Science, Toxicology and Food
8 Technology (IOSR-JESTFT)*, 4(4), 51–58.
- 9 Benjaminsen, T. A. (2016). Does Climate Change Lead to Conflicts in the Sahel? In R. Behnke & M.
10 Mortimore (Eds.), *The End of Desertification? Disputing Environmental Change in the
11 Drylands* (pp. 99–116). Springer, Berlin, Heidelberg. [https://doi.org/10.1007/978-3-642-16014-
12 1_4](https://doi.org/10.1007/978-3-642-16014-1_4)
- 13 Benjaminsen, T. A., Alinon, K., Buhaug, H., & Buset, J. T. (2012). Does climate change drive land-
14 use conflicts in the Sahel? *Journal of Peace Research*, 49(1), 97–111.
15 <https://doi.org/10.1177/0022343311427343>
- 16 Benjaminsen, T. A., & Hiernaux, P. (2019). From Desiccation to Global Climate Change: A History
17 of the Desertification Narrative in the West African Sahel, 1900-2018. *Global Environment*,
18 12(1), 206–236. <https://doi.org/10.3197/ge.2019.120109>
- 19 Benmoussa, H., Luedeling, E., Ghrab, M., Yahmed, J. Ben, & Mimoun, B. (2017). Performance of
20 pistachio (*Pistacia vera* L.) in warming Mediterranean orchards. *Environmental and
21 Experimental Botany*, 140(2017), 76–85. <https://doi.org/10.1016/j.envexpbot.2017.05.007>
- 22 Bennett, C. M., McKendry, I. G., Kelly, S., Denike, K., & Koch, T. (2006). Impact of the 1998 Gobi
23 dust event on hospital admissions in the Lower Fraser Valley, British Columbia. *Science of The
24 Total Environment*, 366(2–3), 918–925. <https://doi.org/10.1016/j.scitotenv.2005.12.025>
- 25 Bensaid, S. (1995). Bilan critique du barrage vert en Algérie. *Science et Changements Planétaires /
26 Sécheresse*, 6(3), 247–255. Retrieved from [http://www.jle.com/fr/revues/sec/e-
27 docs/bilan_critique_du_barrage_vert_en_algerie_270932/article.phtml?tab=citer](http://www.jle.com/fr/revues/sec/e-docs/bilan_critique_du_barrage_vert_en_algerie_270932/article.phtml?tab=citer)
- 28 Bergametti, G., Marticorena, B., Rajot, J. L., Foret, G., Alfaro, S. C., & Laurent, B. (2018). Size-
29 Resolved Dry Deposition Velocities of Dust Particles: In Situ Measurements and
30 Parameterizations Testing. *Journal of Geophysical Research: Atmospheres*, 123(19), 11,080-
31 11,099. <https://doi.org/10.1029/2018JD028964>
- 32 Berman, R., Quinn, C., & Paavola, J. (2012). The role of institutions in the transformation of coping
33 capacity to sustainable adaptive capacity. *Environmental Development*, 2, 86–100.
34 <https://doi.org/10.1016/J.ENVDEV.2012.03.017>
- 35 Berrahmouni, N., Laestadius, L., Martucci, A., Mollicone, D., Patriarca, C., & Sacande, M. (2016).
36 *Building Africa's Great Green Wall Restoring degraded drylands for stronger and more
37 resilient communities*. Rome, Italy. Retrieved from
38 [http://www.greengrowthknowledge.org/sites/default/files/downloads/resource/FAO_Great_Gree
39 n_Wall.pdf](http://www.greengrowthknowledge.org/sites/default/files/downloads/resource/FAO_Great_Green_Wall.pdf)
- 40 Besser, H., Mokadem, N., Redhouania, B., Rhimi, N., Khelifi, F., Ayadi, Y., ... Hamed, Y. (2017).
41 GIS-based evaluation of groundwater quality and estimation of soil salinization and land
42 degradation risks in an arid Mediterranean site (SW Tunisia). *Arabian Journal of Geosciences*,
43 10(16), 350. <https://doi.org/10.1007/s12517-017-3148-0>
- 44 Bestelmeyer, B. T., Ash, A., Brown, J. R., Densambuu, B., Fernández-Giménez, M., Johanson, J., ...
45 Shaver, P. (2017). State and Transition Models: Theory, Applications, and Challenges (pp. 303–
46 345). Springer, Cham. https://doi.org/10.1007/978-3-319-46709-2_9
- 47 Bestelmeyer, B. T., Duniway, M. C., James, D. K., Burkett, L. M., & Havstad, K. M. (2013). A test of
48 critical thresholds and their indicators in a desertification-prone ecosystem: more resilience than

- 1 we thought. *Ecology Letters*, 16(3), 339–345. <https://doi.org/10.1111/ele.12045>
- 2 Bestelmeyer, B. T., Okin, G. S., Duniway, M. C., Archer, S. R., Sayre, N. F., Williamson, J. C., &
3 Herrick, J. E. (2015). Desertification, land use, and the transformation of global drylands.
4 *Frontiers in Ecology and the Environment*, 13(1), 28–36. <https://doi.org/10.1890/140162>
- 5 Bestelmeyer, B. T., Peters, D. P. C., Archer, S. R., Browning, D. M., Okin, G. S., Schooley, R. L., &
6 Webb, N. P. (2018). The Grassland–Shrubland Regime Shift in the Southwestern United States:
7 Misconceptions and Their Implications for Management. *BioScience*, 68(9), 678–690.
8 <https://doi.org/10.1093/biosci/biy065>
- 9 Betts, R. A., Alfieri, L., Bradshaw, C., Caesar, J., Feyen, L., Friedlingstein, P., ... Wyser, K. (2018).
10 Changes in climate extremes, fresh water availability and vulnerability to food insecurity
11 projected at 1.5°C and 2°C global warming with a higher-resolution global climate model.
12 *Philosophical Transactions of the Royal Society A: Mathematical,*
13 *Physical and Engineering Sciences*, 376(2119), 20160452.
14 <https://doi.org/10.1098/rsta.2016.0452>
- 15 Bhalla, R. S., Devi Prasad, K. V., & Pelkey, N. W. (2013). Impact of India’s watershed development
16 programs on biomass productivity. *Water Resources Research*, 49(3), 1568–1580.
17 <https://doi.org/10.1002/wrcr.20133>
- 18 Bhalla, R. S., Pelkey, N. W., & Devi Prasad, K. V. (2011). Application of GIS for Evaluation and
19 Design of Watershed Guidelines. *Water Resources Management*, 25(1), 113–140.
20 <https://doi.org/10.1007/s11269-010-9690-0>
- 21 Bharucha, Z. P., Smith, D., & Pretty, J. (2014). All Paths Lead to Rain: Explaining why Watershed
22 Development in India Does Not Alleviate the Experience of Water Scarcity. *The Journal of*
23 *Development Studies*, 50(9), 1209–1225. <https://doi.org/10.1080/00220388.2014.928699>
- 24 Biazin, B., Sterk, G., Temesgen, M., Abdulkedir, A., & Stroosnijder, L. (2012). Rainwater harvesting
25 and management in rainfed agricultural systems in sub-Saharan Africa – A review. *Physics and*
26 *Chemistry of the Earth, Parts A/B/C*, 47–48, 139–151.
27 <https://doi.org/10.1016/J.PCE.2011.08.015>
- 28 Biederman, J. A., Scott, R. L., Bell, T. W., Bowling, D. R., Dore, S., Garatuza-Payan, J., ... Goulden,
29 M. L. (2017). CO₂ exchange and evapotranspiration across dryland ecosystems of southwestern
30 North America. *Global Change Biology*, 23(10), 4204–4221. <https://doi.org/10.1111/gcb.13686>
- 31 Bikila, N. G., Tessema, Z. K., & Abule, E. G. (2016). Carbon sequestration potentials of semi-arid
32 rangelands under traditional management practices in Borana, Southern Ethiopia. *Agriculture,*
33 *Ecosystems & Environment*, 223, 108–114. <https://doi.org/10.1016/J.AGEE.2016.02.028>
- 34 Bird, M. I., Veenendaal, E. M., Moyo, C., Lloyd, J., & Frost, P. (2000). Effect of fire and soil texture
35 on soil carbon in a sub-humid savanna (Matopos, Zimbabwe). *Geoderma*, 94(1), 71–90.
36 Retrieved from [http://www.sciencedirect.com/science/article/B6V67-3YF3WTR-](http://www.sciencedirect.com/science/article/B6V67-3YF3WTR-5/2/28f7a5524ab958a42f45e1428dbd9806)
37 [5/2/28f7a5524ab958a42f45e1428dbd9806](http://www.sciencedirect.com/science/article/B6V67-3YF3WTR-5/2/28f7a5524ab958a42f45e1428dbd9806)
- 38 Bisigato, A. J., & Laphitz, R. M. L. (2009). Ecohydrological effects of grazing-induced degradation in
39 the Patagonian Monte, Argentina. *Austral Ecology*, 34(5), 545–557.
40 <https://doi.org/10.1111/j.1442-9993.2009.01958.x>
- 41 Bjornlund, H., van Rooyen, A., & Stirzaker, R. (2017). Profitability and productivity barriers and
42 opportunities in small-scale irrigation schemes. *International Journal of Water Resources*
43 *Development*, 33(5), 690–704. <https://doi.org/10.1080/07900627.2016.1263552>
- 44 Bobojonov, I., Lamers, J. P. A., Bekchanov, M., Djanibekov, N., Franz-Vasdeki, J., Ruzimov, J., &
45 Martius, C. (2013). Options and Constraints for Crop Diversification: A Case Study in
46 Sustainable Agriculture in Uzbekistan. *Agroecology and Sustainable Food Systems*, 37(7), 788–
47 811. <https://doi.org/10.1080/21683565.2013.775539>
- 48 Bobojonov, I., Teuber, R., Hasanov, S., Urutyay, V., & Glauben, T. (2016). Farmers’ export market

- 1 participation decisions in transition economies: a comparative study between Armenia and
2 Uzbekistan. *Development Studies Research*, 3(1), 25–35.
3 <https://doi.org/10.1080/21665095.2016.1262272>
- 4 Boelee, E., Yohannes, M., Poda, J.-N., McCartney, M., Cecchi, P., Kibret, S., ... Laamrani, H. (2013).
5 Options for water storage and rainwater harvesting to improve health and resilience against
6 climate change in Africa. *Regional Environmental Change*, 13(3), 509–519.
7 <https://doi.org/10.1007/s10113-012-0287-4>
- 8 Boix-Fayos, C., Martínez-Mena, M., Arnau-Rosalén, E., Calvo-Cases, A., Castillo, V., & Albaladejo,
9 J. (2006). Measuring soil erosion by field plots: Understanding the sources of variation. *Earth-*
10 *Science Reviews*, 78(3–4), 267–285. <https://doi.org/10.1016/J.EARSCIREV.2006.05.005>
- 11 Bond, W. J., & Keeley, J. E. (2005). Fire as a global “herbivore”: the ecology and evolution of
12 flammable ecosystems. *Trends in Ecology and Evolution*, 20(7), 387–394.
- 13 Bond, W. J., & Midgley, G. F. (2000). A proposed CO₂ controlled mechanism of woody plant
14 invasion in grasslands and savannas. *Global Change Biology*, 6, 865–869.
- 15 Bond, W. J., Midgley, G. F., & Woodward, F. I. (2003). What controls South African vegetation -
16 climate or fire? *South African Journal of Botany*, 69, 79–91.
- 17 Bonilla, C. A., Reyes, J. L., & Magri, A. (2010). Water Erosion Prediction Using the Revised
18 Universal Soil Loss Equation (RUSLE) in a GIS Framework, Central Chile. *Chilean Journal of*
19 *Agricultural Research*, 70(1), 159–169. <https://doi.org/10.4067/S0718-58392010000100017>
- 20 Bonney, R., Phillips, T. B., Ballard, H. L., & Enck, J. W. (2016). Can citizen science enhance public
21 understanding of science? *Public Understanding of Science*, 25(1), 2–16.
22 <https://doi.org/10.1177/0963662515607406>
- 23 Booker, J. F., Michelsen, A. M., & Ward, F. A. (2005). Economic impact of alternative policy
24 responses to prolonged and severe drought in the Rio Grande Basin. *Water Resources Research*,
25 41(2). <https://doi.org/10.1029/2004WR003486>
- 26 Booth, B. B. B., Dunstone, N. J., Halloran, P. R., Andrews, T., & Bellouin, N. (2012). Aerosols
27 implicated as a prime driver of twentieth-century North Atlantic climate variability. *Nature*,
28 484(7393), 228–232. <https://doi.org/10.1038/nature10946>
- 29 Booyesen, M. J., Visser, M., & Burger, R. (2019). Temporal case study of household behavioural
30 response to Cape Town’s “Day Zero” using smart meter data. *Water Research*, 149, 414–420.
31 <https://doi.org/10.1016/J.WATRES.2018.11.035>
- 32 Börner, J., Baylis, K., Corbera, E., Ezzine-de-Blas, D., Honey-Rosés, J., Persson, U. M., & Wunder,
33 S. (2017). The Effectiveness of Payments for Environmental Services. *World Development*, 96,
34 359–374. <https://doi.org/10.1016/J.WORLDDEV.2017.03.020>
- 35 Borrelli, P., Robinson, D. A., Fleischer, L. R., Lugato, E., Ballabio, C., Alewell, C., ... Panagos, P.
36 (2017). An assessment of the global impact of 21st century land use change on soil erosion.
37 *Nature Communications*, 8(1), 2013. <https://doi.org/10.1038/s41467-017-02142-7>
- 38 Bose, P. (n.d.). India’s drylands agroforestry: a ten-year analysis of gender and social diversity, tenure
39 and climate variability. *International Forestry Review*, 17(S4). Retrieved from
40 <http://www.ingentaconnect.com/content/cfa/ifr/2015/00000017/A00404s4/art00008?crawler=true>
41 e
- 42 Boserup, E. (1965). *The conditions of agricultural growth: the economics of agrarian change under*
43 *population pressure Earthscan*. Oxford: Earthscan.
- 44 Botterill, L. C., & Hayes, M. J. (2012). Drought triggers and declarations: Science and policy
45 considerations for drought risk management. *Natural Hazards*, 64(1), 139–151.
46 <https://doi.org/10.1007/s11069-012-0231-4>
- 47 Boucher, O., Randall, D., Artaxo, P., Bretherton, C., Feingold, G., Forster, P., ... Zhang, X. Y.

- 1 (2013). Clouds and Aerosols. In T. F. Stocker, D. Qin, G.-K. Plattner, M. Tignor, S. K. Allen, J.
2 Boschung, ... P. M. Midgley (Eds.), *Climate Change 2013: The Physical Science Basis.*
3 *Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel*
4 *on Climate Change* (pp. 571–657). Cambridge, United Kingdom and New York, NY, USA:
5 Cambridge University Press.
- 6 Bouma, J., Van Soest, D., & Bulte, E. (2007). How sustainable is participatory watershed
7 development in India? *Agricultural Economics*, *36*(1), 13–22.
- 8 Bouma, J., & Wösten, J. H. M. (2016). How to characterize ‘good’ and ‘greening’ in the EU Common
9 Agricultural Policy (CAP): the case of clay soils in the Netherlands. *Soil Use and Management*,
10 *32*(4), 546–552. <https://doi.org/10.1111/sum.12289>
- 11 Bourguignon, F., & Chakravarty, S. R. (2003). The Measurement of Multidimensional Poverty. *The*
12 *Journal of Economic Inequality*, *1*(1), 25–49. <https://doi.org/10.1023/A:1023913831342>
- 13 Boyd, P. W., Mackie, D. S., & Hunter, K. A. (2010). Aerosol iron deposition to the surface ocean —
14 Modes of iron supply and biological responses. *Marine Chemistry*, *120*(1–4), 128–143.
15 <https://doi.org/10.1016/J.MARCHEM.2009.01.008>
- 16 Bradley, B. A., Blumenthal, D. M., Early, R., Grosholz, E. D., Lawler, J. J., Miller, L. P., ... Olden, J.
17 D. (2012). Global change, global trade, and the next wave of plant invasions. *Frontiers in*
18 *Ecology and the Environment*, *10*(1), 20–28. <https://doi.org/10.1890/110145>
- 19 Bradley, B. A., Blumenthal, D. M., Wilcove, D. S., & Ziska, L. H. (2010). Predicting plant invasions
20 in an era of global change. *Trends in Ecology & Evolution*, *25*(5), 310–318.
21 <https://doi.org/10.1016/J.TREE.2009.12.003>
- 22 Brandt, M., Hiernaux, P., Rasmussen, K., Mbow, C., Kergoat, L., Tagesson, T., ... Fensholt, R.
23 (2016). Assessing woody vegetation trends in Sahelian drylands using MODIS based seasonal
24 metrics. *Remote Sensing of Environment*, *183*, 215–225.
25 <https://doi.org/10.1016/J.RSE.2016.05.027>
- 26 Brandt, M., Hiernaux, P., Tagesson, T., Verger, A., Rasmussen, K., Diouf, A. A., ... Fensholt, R.
27 (2016). Woody plant cover estimation in drylands from Earth Observation based seasonal
28 metrics. *Remote Sensing of Environment*, *172*, 28–38.
29 <https://doi.org/10.1016/J.RSE.2015.10.036>
- 30 Brandt, M., Mbow, C., Diouf, A. A., Verger, A., Samimi, C., & Fensholt, R. (2015). Ground- and
31 satellite-based evidence of the biophysical mechanisms behind the greening Sahel. *Global*
32 *Change Biology*, *21*(4), 1610–1620. <https://doi.org/10.1111/gcb.12807>
- 33 Breckle, S.-W. (2013). From Aral Sea to Aralkum: An Ecological Disaster or Halophytes’ Paradise
34 (pp. 351–398). Springer, Berlin, Heidelberg. https://doi.org/10.1007/978-3-642-30967-0_13
- 35 Broeckhoven, N., & Cliquet, A. (2015). Gender and ecological restoration: time to connect the dots.
36 *Restoration Ecology*, *23*(6), 729–736. <https://doi.org/10.1111/rec.12270>
- 37 Brooks, M. L., D’Antonio, C. M., Richardson, D. M., Grace, J. B., Keeley, J. E., DiTomaso, J. M., ...
38 Pyke, D. (2004). Effects of Invasive Alien Plants on Fire Regimes. *BioScience*, *54*(7), 677–688.
39 [https://doi.org/10.1641/0006-3568\(2004\)054\[0677:eoiapo\]2.0.co;2](https://doi.org/10.1641/0006-3568(2004)054[0677:eoiapo]2.0.co;2)
- 40 Browning, D. M., & Archer, S. R. (2011). Protection from livestock fails to deter shrub proliferation
41 in a desert landscape with a history of heavy grazing. *Ecological Applications*, *21*(5), 1629–
42 1642. <https://doi.org/10.1890/10-0542.1>
- 43 Browning, D. M., Maynard, J. J., Karl, J. W., & Peters, D. C. (2017). Breaks in MODIS time series
44 portend vegetation change: verification using long-term data in an arid grassland ecosystem.
45 *Ecological Applications*, *27*(5), 1677–1693. <https://doi.org/10.1002/eap.1561>
- 46 Bruno, L., Horvat, M., & Raffaele, L. (2018). Windblown sand along railway infrastructures: A
47 review of challenges and mitigation measures. *Journal of Wind Engineering and Industrial*
48 *Aerodynamics*, *177*, 340–365. <https://doi.org/10.1016/J.JWEIA.2018.04.021>

- 1 Bryan, E., Deressa, T., & ... G. G. (2009). Adaptation to climate change in Ethiopia and South Africa:
2 options and constraints. *Environmental Science & Policy*, 12(4). Retrieved from
3 <http://www.sciencedirect.com/science/article/pii/S1462901108001263>
- 4 Buhaug, H., Nordkvelle, J., Bernauer, T., Böhmelt, T., Brzoska, M., Busby, J. W., ... von Uexkull, N.
5 (2014). One effect to rule them all? A comment on climate and conflict. *Climatic Change*,
6 127(3–4), 391–397. <https://doi.org/10.1007/s10584-014-1266-1>
- 7 Bulkeley, H. (2013). *Cities and Climate Change*. Routledge. <https://doi.org/10.4324/9780203077207>
- 8 Bureau of Land Management. (2005). *Vegetation Treatments on Bureau of Land Management Lands*
9 *in 17 Western State*. Retrieved from
10 <https://digitalcommons.usu.edu/cgi/viewcontent.cgi?referer=https://www.google.com.my/&https://doi.org/10.1016/J.WORLDDEV.2014.10.034>
11 <https://doi.org/10.1016/J.WORLDDEV.2014.10.034>
- 12 Burke, P. J., & Dundas, G. (2015). Female Labor Force Participation and Household Dependence on
13 Biomass Energy: Evidence from National Longitudinal Data. *World Development*, 67, 424–437.
14 <https://doi.org/10.1016/J.WORLDDEV.2014.10.034>
- 15 Burney, J., Woltering, L., Burke, M., Naylor, R., & Pasternak, D. (2010). Solar-powered drip
16 irrigation enhances food security in the Sudano-Sahel. *Proceedings of the National Academy of*
17 *Sciences of the United States of America*, 107(5), 1848–53.
18 <https://doi.org/10.1073/pnas.0909678107>
- 19 Burrell, A. L., Evans, J. P., & Liu, Y. (2017). Detecting dryland degradation using Time Series
20 Segmentation and Residual Trend analysis (TSS-RESTREND). *Remote Sensing of Environment*,
21 197. <https://doi.org/10.1016/j.rse.2017.05.018>
- 22 Burrell, A. L., Evans, J. P., & Liu, Y. (2018). The impact of dataset selection on land degradation
23 assessment. *ISPRS Journal of Photogrammetry and Remote Sensing*, 146, 22–37.
24 <https://doi.org/10.1016/J.ISPRSJPRS.2018.08.017>
- 25 Byers, E., Gidden, M., Leclère, D., Balkovic, J., Burek, P., Ebi, K., ... Riahi, K. (2018). Global
26 exposure and vulnerability to multi-sector development and climate change hotspots.
27 *Environmental Research Letters*, 13(5), 055012. <https://doi.org/10.1088/1748-9326/aabf45>
- 28 Cai, X., McKinney, D. C., & Rosegrant, M. W. (2003). Sustainability analysis for irrigation water
29 management in the Aral Sea region. *Agricultural Systems*, 76(3), 1043–1066.
30 [https://doi.org/10.1016/S0308-521X\(02\)00028-8](https://doi.org/10.1016/S0308-521X(02)00028-8)
- 31 Cai, X., Zhang, X., & Wang, D. (2011). Land Availability for Biofuel Production. *Environmental*
32 *Science & Technology*, 45(1), 334–339. <https://doi.org/10.1021/es103338e>
- 33 Calder, I., Gosain, A., Rao, M. S. R. M., Batchelor, C., Snehalatha, M., & Bishop, E. (2008).
34 Watershed development in India. 1. Biophysical and societal impacts. *Environment*,
35 *Development and Sustainability*, 10(4), 537–557. <https://doi.org/10.1007/s10668-006-9079-7>
- 36 Çalişkan, S., & Boydak, M. (2017). Afforestation of arid and semiarid ecosystems in Turkey.
37 *TURKISH JOURNAL OF AGRICULTURE AND FORESTRY*, 41, 317–330.
38 <https://doi.org/10.3906/tar-1702-39>
- 39 Calvet-Mir, L., Corbera, E., Martin, A., Fisher, J., & Gross-Camp, N. (2015). Payments for ecosystem
40 services in the tropics: a closer look at effectiveness and equity. *Current Opinion in*
41 *Environmental Sustainability*, 14, 150–162. <https://doi.org/10.1016/J.COSUST.2015.06.001>
- 42 Cancino-Solórzano, Y., Paredes-Sánchez, J. P., Gutiérrez-Trashorras, A. J., & Xiberta-Bernat, J.
43 (2016). The development of renewable energy resources in the State of Veracruz, Mexico.
44 *Utilities Policy*, 39, 1–4. <https://doi.org/10.1016/J.JUP.2016.01.001>
- 45 Cano, A., Núñez, A., Acosta-Martinez, V., Schipanski, M., Ghimire, R., Rice, C., & West, C. (2018).
46 Current knowledge and future research directions to link soil health and water conservation in
47 the Ogallala Aquifer region. *Geoderma*, 328, 109–118.
48 <https://doi.org/10.1016/J.GEODERMA.2018.04.027>

- 1 Cao, J., Zhang, X., Deo, R., Gong, Y., & Feng, Q. (2018). Influence of stand type and stand age on
2 soil carbon storage in China's arid and semi-arid regions. *Land Use Policy*, 78, 258–265.
3 <https://doi.org/10.1016/J.LANDUSEPOL.2018.07.002>
- 4 Çarkacı, D. A. (1999). *EXAMPLES OF GOOD PRACTICES OF TURKEY FOR COMBATING WIND*
5 *EROSION (EXAMPLES OF KARAPINAR) THE BLOOMING DESERT KARAPINAR*
6 *SITUATION OF DESERTIFICATION IN TURKEY Distribution of Problem Areas in Turkey.*
- 7 Carletto, C., Corral, P., & Guelfi, A. (2017). Agricultural commercialization and nutrition revisited:
8 Empirical evidence from three African countries. *Food Policy*, 67, 106–118.
9 <https://doi.org/10.1016/J.FOODPOL.2016.09.020>
- 10 Carlisle, K., & Gruby, R. L. (2017). Polycentric Systems of Governance: A Theoretical Model for the
11 Commons. *Policy Studies Journal*. <https://doi.org/10.1111/psj.12212>
- 12 Carr-Hill, R. (2013). Missing Millions and Measuring Development Progress. *World Development*,
13 46, 30–44. <https://doi.org/10.1016/J.WORLDDEV.2012.12.017>
- 14 Carr, E. R., & Thompson, M. C. (2014). Gender and Climate Change Adaptation in Agrarian Settings:
15 Current Thinking, New Directions, and Research Frontiers. *Geography Compass*, 8(3), 182–197.
16 <https://doi.org/10.1111/gec3.12121>
- 17 Carrão, H., Naumann, G., & Barbosa, P. (2017). Global projections of drought hazard in a warming
18 climate: a prime for disaster risk management. *Climate Dynamics*.
19 <https://doi.org/10.1007/s00382-017-3740-8>
- 20 Cassidy, T. M., Fownes, J. H., & Harrington, R. A. (2004). Nitrogen Limits an Invasive Perennial
21 Shrub in Forest Understory. *Biological Invasions*, 6(1), 113–121.
22 <https://doi.org/10.1023/B:BINV.0000010128.44332.0f>
- 23 Castellanos-Villegas, A. E., Yanes, G., & Valdes, D. (2002). Drought-tolerant exotic buffel-grass and
24 desertification. In *Weeds across borders: proceedings of a North American conference held at*
25 *the Arizona-Sonora desert museum*. Tucson.
- 26 Castellanos, A. E., Celaya-Michel, H., Rodríguez, J. C., & Wilcox, B. P. (2016). Ecohydrological
27 changes in semiarid ecosystems transformed from shrubland to buffelgrass savanna.
28 *Ecohydrology*, 9(8), 1663–1674. <https://doi.org/10.1002/eco.1756>
- 29 Catacutan, D. C., & Villamor, G. B. (2016). Gender Roles and Land Use Preferences—Implications
30 to Landscape Restoration in Southeast Asia. In *Land Restoration* (pp. 431–440). Elsevier.
31 <https://doi.org/10.1016/B978-0-12-801231-4.00029-X>
- 32 Cattaneo, C., & Peri, G. (2016). The migration response to increasing temperatures. *Journal of*
33 *Development Economics*, 122, 127–146. <https://doi.org/10.1016/J.JDEVECO.2016.05.004>
- 34 Ceballos, G., Pacheco, J., Cruzado, J., Manzano-Fischer, P., List, R., Davidson, A., & Santos-Barrera,
35 G. (2010). Rapid Decline of a Grassland System and Its Ecological and Conservation
36 Implications. *PLoS ONE*, 5(1), e8562. <https://doi.org/10.1371/journal.pone.0008562>
- 37 CGIAR-RPDS. (2014). *Annual report 2 0 1 4. Pathways to lasting impact for rural dryland*
38 *communities in the developing world*. Amman, Jordan. Retrieved from
39 http://drylandsystems.cgiar.org/sites/default/files/DS_annual_report_2014.pdf
- 40 Chambers, J. C., Bradley, B. A., Brown, C. S., D'Antonio, C., Germino, M. J., Grace, J. B., ... Pyke,
41 D. A. (2014). Resilience to Stress and Disturbance, and Resistance to *Bromus tectorum* L.
42 Invasion in Cold Desert Shrublands of Western North America. *Ecosystems*, 17(2), 360–375.
43 <https://doi.org/10.1007/s10021-013-9725-5>
- 44 Chandel, S. S., Nagaraju Naik, M., & Chandel, R. (2015). Review of solar photovoltaic water
45 pumping system technology for irrigation and community drinking water supplies. *Renewable*
46 *and Sustainable Energy Reviews*, 49, 1084–1099. <https://doi.org/10.1016/J.RSER.2015.04.083>
- 47 Chandrasekhar, C. P., Ghosh, J., & Roychowdhury, A. (2006). The 'demographic dividend' and young

- 1 India's economic future. *Economic and Political Weekly*, 5055–5064.
- 2 Charney, J., Stone, P. H., & Quirk, W. J. (1975). Drought in the sahara: a biogeophysical feedback
3 mechanism. *Science (New York, N.Y.)*, 187(4175), 434–5.
4 <https://doi.org/10.1126/science.187.4175.434>
- 5 Chasek, P., Akhtar-Schuster, M., Orr, B. J., Luise, A., Rakoto Ratsimba, H., & Safriel, U. (2019).
6 Land degradation neutrality: The science-policy interface from the UNCCD to national
7 implementation. *Environmental Science & Policy*, 92, 182–190.
8 <https://doi.org/10.1016/j.envsci.2018.11.017>
- 9 Chasek, P., Essahli, W., Akhtar-Schuster, M., Stringer, L. C., & Thomas, R. (2011). Integrated land
10 degradation monitoring and assessment: Horizontal knowledge management at the national and
11 international levels. *Land Degradation & Development*, 22(2), 272–284.
12 <https://doi.org/10.1002/ldr.1096>
- 13 Cheesman, S., Thierfelder, C., Eash, N. S., Kassie, G. T., & Frossard, E. (2016). Soil carbon stocks in
14 conservation agriculture systems of Southern Africa. *Soil and Tillage Research*, 156, 99–109.
15 <https://doi.org/https://doi.org/10.1016/j.still.2015.09.018>
- 16 Chen, D., Mi, J., Chu, P., Cheng, J., Zhang, L., Pan, Q., ... Bai, Y. (2015). Patterns and drivers of soil
17 microbial communities along a precipitation gradient on the Mongolian Plateau. *Landscape*
18 *Ecology*, 30(9), 1669–1682. <https://doi.org/10.1007/s10980-014-9996-z>
- 19 Chen, J., Yin, S., Gebhardt, H., & Yang, X. (2018). Farmers' livelihood adaptation to environmental
20 change in an arid region: A case study of the Minqin Oasis, northwestern China. *Ecological*
21 *Indicators*, 93, 411–423. <https://doi.org/10.1016/J.ECOLIND.2018.05.017>
- 22 Chen, R., Ye, C., Cai, Y., Xing, X., & Chen, Q. (2014). The impact of rural out-migration on land use
23 transition in China: Past, present and trend. *Land Use Policy*, 40, 101–110.
24 <https://doi.org/10.1016/J.LANDUSEPOL.2013.10.003>
- 25 Chen, S., Lin, G., H, J., & Jenerette, G. D. (2009). Dependence of carbon sequestration on the
26 differential responses of ecosystem photosynthesis and respiration to rain pulses in a semiarid
27 steppe. *Global Change Biology*, 15(10), 2450–2461. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2486.2009.01879.x)
28 [2486.2009.01879.x](https://doi.org/10.1111/j.1365-2486.2009.01879.x)
- 29 Chen, T., Bao, A., Jiapaer, G., Guo, H., Zheng, G., Jiang, L., ... Tuerhanjiang, L. (2019).
30 Disentangling the relative impacts of climate change and human activities on arid and semiarid
31 grasslands in Central Asia during 1982–2015. *Science of The Total Environment*, 653, 1311–
32 1325. <https://doi.org/10.1016/j.scitotenv.2018.11.058>
- 33 Chen, Y., Li, Z., Fan, Y., Wang, H., & Deng, H. (2015). Progress and prospects of climate change
34 impacts on hydrology in the arid region of northwest China. *Environmental Research*, 139, 11–
35 19. <https://doi.org/10.1016/J.ENVRES.2014.12.029>
- 36 Cheng, J., & Xue, C. (2014). The sand-damage–prevention engineering system for the railway in the
37 desert region of the Qinghai-Tibet plateau. *Journal of Wind Engineering and Industrial*
38 *Aerodynamics*, 125, 30–37. <https://doi.org/10.1016/J.JWEIA.2013.11.016>
- 39 Chengrui, M., & Dregne, H. E. (2001). Review article: Silt and the future development of China's
40 Yellow River. *The Geographical Journal*, 167(1), 7–22. [https://doi.org/10.1111/1475-](https://doi.org/10.1111/1475-4959.00002)
41 [4959.00002](https://doi.org/10.1111/1475-4959.00002)
- 42 Cherlet, M., Hutchinson, C., Reynolds, J., Hill, J., Sommer, S., & von Maltitz, G. (2018). *World Atlas*
43 *of Desertification*. (M. Cherlet, C. Hutchinson, J. Reynolds, J. Hill, S. Sommer, & G. von
44 Maltitz, Eds.) (3rd ed.). Luxembourg: Publication Office of the European Union.
45 <https://doi.org/10.2760/9205>
- 46 Choobari, O. A., Zavar-Reza, P., & Sturman, A. (2014). The global distribution of mineral dust and
47 its impacts on the climate system: A review. *Atmospheric Research*, 138, 152–165.
48 <https://doi.org/10.1016/J.ATMOSRES.2013.11.007>

- 1 Chowdhury, S., & Al-Zahrani, M. (2013). Implications of Climate Change on Water Resources in
2 Saudi Arabia. *Arabian Journal for Science and Engineering*, 38(8), 1959–1971.
3 <https://doi.org/10.1007/s13369-013-0565-6>
- 4 Christian, B. A., Dhinwa, P. S., & Ajai. (2018). Long term monitoring and assessment of
5 desertification processes using medium and high resolution satellite data. *Applied Geography*,
6 97, 10–24. <https://doi.org/10.1016/J.APGEOG.2018.04.010>
- 7 Christie, F., & Hanlon, J. (2001). *Mozambique and the great flood of 2000*. Bloomington and
8 Indianapolis: Indiana University Press.
- 9 Christy, R. D., da Silva, C. A., Mhlanga, N., Mabaya, E., & Tihanyi, K. (2014). *Innovative*
10 *Institutions, Public Policies and Private Strategies for Agro-Enterprise Development*. CO-
11 PUBLISH WITH FOOD AND AGRICULTURE ORGANIZATION OF THE UNITED
12 NATIONS (FAO). <https://doi.org/10.1142/9131>
- 13 Chu, X., Zhan, J., Li, Z., Zhang, F., & Qi, W. (2019). Assessment on forest carbon sequestration in
14 the Three-North Shelterbelt Program region, China. *Journal of Cleaner Production*, 215, 382–
15 389. <https://doi.org/10.1016/J.JCLEPRO.2018.12.296>
- 16 Chughtai, M. I., Mahmood, K., & Awan, A. R. (2015). *Growth Performance of Carp Species Fed on*
17 *Salt-Tolerant Roughages and Formulated Feed in Brackish Water Under Polyculture System*.
18 *Pakistan Journal of Zoology* (Vol. 47). Retrieved from <http://www.zsp.com.pk/pdf47/775-781>
19 (23) PJZ-688-11 3-4-15 Corrected.pdf
- 20 Chytrý, M., Wild, J., Pyšek, P., Jarošík, V., Dendoncker, N., Reginster, I., ... Settele, J. (2012).
21 Projecting trends in plant invasions in Europe under different scenarios of future land-use
22 change. *Global Ecology and Biogeography*, 21(1), 75–87. [https://doi.org/10.1111/j.1466-
23 8238.2010.00573.x](https://doi.org/10.1111/j.1466-)
- 24 CIREN. (2010). *Determinación de la erosión potencial y actual de los suelos de Chile*. Centro de
25 Información de Recursos Naturales.
- 26 Clarke, D. J., & Hill, R. V. (2013). *Cost-benefit analysis of the African risk capacity facility* (No.
27 01292). Washington D.C. Retrieved from
28 <http://ebrary.ifpri.org/cdm/ref/collection/p15738coll2/id/127813>
- 29 Clarke, H., & Evans, J. P. (2018). Exploring the future change space for fire weather in southeast
30 Australia. *Theoretical and Applied Climatology*. <https://doi.org/10.1007/s00704-018-2507-4>
- 31 Closas, A., & Rap, E. (2017). Solar-based groundwater pumping for irrigation: Sustainability,
32 policies, and limitations. *Energy Policy*, 104, 33–37.
33 <https://doi.org/10.1016/J.ENPOL.2017.01.035>
- 34 Codjoe, S. N. A., Owusu, G., & Burkett, V. (2014). Perception, experience, and indigenous
35 knowledge of climate change and variability: the case of Accra, a sub-Saharan African city.
36 *Regional Environmental Change*, 14(1), 369–383. <https://doi.org/10.1007/s10113-013-0500-0>
- 37 Collins, J. A., Prange, M., Caley, T., Gimeno, L., Beckmann, B., Mulitza, S., ... Schefuß, E. (2017).
38 Rapid termination of the African Humid Period triggered by northern high-latitude cooling.
39 *Nature Communications*, 8(1), 1372. <https://doi.org/10.1038/s41467-017-01454-y>
- 40 Conant, R. T., & Paustian, K. (2003). Potential soil carbon sequestration in overgrazed grassland
41 ecosystems. *Global Biogeochemical Cycles*, 16(4), 90-1-90–9.
42 <https://doi.org/10.1029/2001gb001661>
- 43 Coninck, H. de, Revi, A., Babiker, M., Bertoldi, P., Buckeridge, M., Cartwright, A., ... Sugiyama, T.
44 (2018). Chapter 4: Strengthening and implementing the global response. In *IPCC Special Report*
45 *on Global Warming of 1.5 °C*.
- 46 Cook, B. I., Miller, R. L., & Seager, R. (2009). Amplification of the North American “Dust Bowl”
47 drought through human-induced land degradation. *Proceedings of the National Academy of*
48 *Sciences of the United States of America*, 106(13), 4997–5001.

- 1 <https://doi.org/10.1073/pnas.0810200106>
- 2 Cook, B. I., Smerdon, J. E., Seager, R., & Coats, S. (2014a). Global warming and 21st century drying.
3 *Climate Dynamics*, 43(9–10), 2607–2627. <https://doi.org/10.1007/s00382-014-2075-y>
- 4 Cook, B. I., Smerdon, J. E., Seager, R., Cook, E. R., Cook, B. I., Smerdon, J. E., ... Cook, E. R.
5 (2014b). Pan-Continental Droughts in North America over the Last Millennium. *Journal of*
6 *Climate*, 27(1), 383–397. <https://doi.org/10.1175/JCLI-D-13-00100.1>
- 7 Cook, B., Zeng, N., & Yoon, J. H. (2012). Will Amazonia dry out? Magnitude and causes of change
8 from IPCC climate model projections. *Earth Interactions*, 16(3).
9 <https://doi.org/10.1175/2011EI398.1>
- 10 Coppola, E., & Giorgi, F. (2009). An assessment of temperature and precipitation change projections
11 over Italy from recent global and regional climate model simulations. *International Journal of*
12 *Climatology*, 30(1), n/a-n/a. <https://doi.org/10.1002/joc.1867>
- 13 Cordingley, J. E., Snyder, K. A., Rosendahl, J., Kizito, F., & Bossio, D. (2015). Thinking outside the
14 plot: addressing low adoption of sustainable land management in sub-Saharan Africa. *Current*
15 *Opinion in Environmental Sustainability*, 15, 35–40.
16 <https://doi.org/10.1016/J.COSUST.2015.07.010>
- 17 Cornell, S., Berkhout, F., Tuinstra, W., Tàbara, J. D., Chabay, I., de Wit, B., ... van Kerkhoff, L.
18 (2013). Opening up knowledge systems for better responses to global environmental change.
19 *Environmental Science & Policy*, 28, 60–70. <https://doi.org/10.1016/J.ENVSCI.2012.11.008>
- 20 Cornet, A. (2012). Surveillance environnementale et développement Des observations écologiques à
21 la surveillance environnementale: un besoin pour comprendre et agir. *Options*
22 *Méditerranéennes*, B(68), 1–14. Retrieved from
23 http://www.iamm.ciheam.org/ress_doc/opac_css/doc_num.php?explnum_id=7624
- 24 Costa, S. C. S., Diniz, A. S. A. C., & Kazmerski, L. L. (2016). Dust and soiling issues and impacts
25 relating to solar energy systems: Literature review update for 2012–2015. *Renewable and*
26 *Sustainable Energy Reviews*, 63, 33–61. <https://doi.org/10.1016/J.RSER.2016.04.059>
- 27 Costa, S. C. S., Diniz, A. S. A. C., & Kazmerski, L. L. (2018). Solar energy dust and soiling
28 R&D progress: Literature review update for 2016. *Renewable and Sustainable Energy*
29 *Reviews*, 82, 2504–2536. <https://doi.org/10.1016/J.RSER.2017.09.015>
- 30 Costanza, R., Groot, R. de, & P Sutton. (2014). Changes in the global value of ecosystem services.
31 *Global Environmental Change*, 26, 152–158. Retrieved from
32 <http://www.sciencedirect.com/science/article/pii/S0959378014000685>
- 33 Cour, J. M. (2001). The Sahel in West Africa: countries in transition to a full market economy. *Global*
34 *Environmental Change*, 11(1), 31–47. [https://doi.org/10.1016/S0959-3780\(00\)00043-1](https://doi.org/10.1016/S0959-3780(00)00043-1)
- 35 Cowie, A. L., Orr, B. J., Castillo Sanchez, V. M., Chasek, P., Crossman, N. D., Erlewein, A., ...
36 Welton, S. (2018). Land in balance: The scientific conceptual framework for Land Degradation
37 Neutrality. *Environmental Science & Policy*, 79, 25–35.
38 <https://doi.org/10.1016/J.ENVSCI.2017.10.011>
- 39 Cox, J. R., Martin-R, M. H., Ibarra-F, F. A., Fourie, J. H., Rethman, J. F. G., & Wilcox, D. G. (1988).
40 The Influence of Climate and Soils on the Distribution of Four African Grasses. *Journal of*
41 *Range Management*, 41(2), 127. <https://doi.org/10.2307/3898948>
- 42 Crawford, J.A., Olson, R.A., West, N.E., Mosley, J.C., Schroeder, M.A., Whitson, T.D., Miller, R.F.,
43 Gregg, M.A., Boyd, C. S. (2004). Ecology and management of sage-grouse and sage-grouse
44 habitat. *Journal of Range Management*, 57(1), 2–19. [https://doi.org/10.2111/1551-5028\(2004\)057\[0002:EAMOS\]2.0.CO;2](https://doi.org/10.2111/1551-5028(2004)057[0002:EAMOS]2.0.CO;2)
- 46 CRED. (2015). *Human Cost of Natural Disasters 2015: A global perspective*. Retrieved from
47 https://www.preventionweb.net/files/42895_cerdthehumancostofdisastersglobalpe.pdf

- 1 Crovetto, C. . (1998). No-till development in Chequén Farm and its influence on some physical,
2 chemical and biological parameters. *Journal of Soil and Water Conservation*, 58(3), 194–199.
- 3 Cullman, G. (2015). Community Forest Management as Virtualism in Northeastern Madagascar.
4 *Human Ecology*, 43(1), 29–41. <https://doi.org/10.1007/s10745-015-9725-5>
- 5 D’Odorico, P., Bhattachan, A., Davis, K. F., Ravi, S., Runyan, C. W., D’Odorico, P., ... Runyan, C.
6 W. (2013). Global desertification: Drivers and feedbacks. *Advances in Water Resources*,
7 51(October 2017), 326–344. <https://doi.org/10.1016/j.advwatres.2012.01.013>
- 8 Dagar, J. C., Pandey, C. B., & Chaturvedi, C. S. (2014). Agroforestry: a way forward for sustaining
9 fragile coastal and island agro-ecosystems. In J. . Dagar, A. K. Singh, & A. Arunachalam (Eds.),
10 *Agroforestry systems in India: livelihood security & environmental services advances in*
11 *agroforestry. Vol. 10* (pp. 185–232). Springer, New Delhi. [https://doi.org/10.1007/978-81-322-](https://doi.org/10.1007/978-81-322-1662-9_7)
12 [1662-9_7](https://doi.org/10.1007/978-81-322-1662-9_7)
- 13 Dah-gbeto, A. P., & Villamor, G. B. (2016). Gender-specific responses to climate variability in a
14 semi-arid ecosystem in northern Benin. *Ambio*, 45(S3), 297–308.
15 <https://doi.org/10.1007/s13280-016-0830-5>
- 16 Dahiya, B. (2012). Cities in Asia, 2012: Demographics, economics, poverty, environment and
17 governance. *Cities*, 29, S44–S61. <https://doi.org/10.1016/J.CITIES.2012.06.013>
- 18 Dai, A. (2011). Drought under global warming: A review. *Wiley Interdisciplinary Reviews: Climate*
19 *Change*, 2(1), 45–65. <https://doi.org/10.1002/wcc.81>
- 20 Dai, A. (2013). Increasing drought under global warming in observations and models. *Nature Climate*
21 *Change*, 3(1), 52–58. <https://doi.org/10.1038/nclimate1633>
- 22 Damberg, L., & AghaKouchak, A. (2014). Global trends and patterns of drought from space.
23 *Theoretical and Applied Climatology*, 117(3–4), 441–448. [https://doi.org/10.1007/s00704-013-](https://doi.org/10.1007/s00704-013-1019-5)
24 [1019-5](https://doi.org/10.1007/s00704-013-1019-5)
- 25 Das, R., Evan, A., & Lawrence, D. (2013). Contributions of long-distance dust transport to
26 atmospheric P inputs in the Yucatan Peninsula. *Global Biogeochemical Cycles*, 27(1), 167–175.
27 <https://doi.org/10.1029/2012GB004420>
- 28 Davies, G. M., Bakker, J. D., Dettweiler-Robinson, E., Dunwiddie, P. W., Hall, S. A., Downs, J., &
29 Evans, J. (2012). Trajectories of change in sagebrush steppe vegetation communities in relation
30 to multiple wildfires. *Ecological Applications*, 22(5), 1562–1577. [https://doi.org/10.1890/10-](https://doi.org/10.1890/10-2089.1)
31 [2089.1](https://doi.org/10.1890/10-2089.1)
- 32 Davies, J. (2017). *The Land in Drylands: Thriving In Uncertainty through Diversity* (Global Land
33 Outlook Working Paper). Retrieved from
34 [https://static1.squarespace.com/static/5694c48bd82d5e9597570999/t/593660f915d5dbd832a36e](https://static1.squarespace.com/static/5694c48bd82d5e9597570999/t/593660f915d5dbd832a36e5b/1496735996732/The+Land+in+Drylands__J_Davies.pdf)
35 [5b/1496735996732/The+Land+in+Drylands__J_Davies.pdf](https://static1.squarespace.com/static/5694c48bd82d5e9597570999/t/593660f915d5dbd832a36e5b/1496735996732/The+Land+in+Drylands__J_Davies.pdf)
- 36 Davies, J., Herrera, P., Ruiz-Mirazo, J., Mohamed-Katerere, J., Hannam, I., & Nuesiri, E. (2016).
37 *Improving governance of pastoral lands: Implementing the voluntary guidelines on the*
38 *responsible governance of tenure of land, fisheries and forests in the context of national food*
39 *security*.
- 40 Davis, D. K. (2016). *The arid lands : history, power, knowledge*. MIT Press. Retrieved from
41 <https://mitpress.mit.edu/books/arid-lands>
- 42 Davis, M. A., Grime, J. P., & Thompson, K. (2000). Fluctuating resources in plant communities: a
43 general theory of invasibility. *Journal of Ecology*, 88(3), 528–534.
44 <https://doi.org/10.1046/j.1365-2745.2000.00473.x>
- 45 Dawelbait, M., & Morari, F. (2012). Monitoring desertification in a Savannah region in Sudan using
46 Landsat images and spectral mixture analysis. *Journal of Arid Environments*, 80, 45–55.
47 <https://doi.org/10.1016/j.jaridenv.2011.12.011>

- 1 de Châtel, F. (2014). The Role of Drought and Climate Change in the Syrian Uprising: Untangling the
2 Triggers of the Revolution. *Middle Eastern Studies*, 50(4), 521–535.
3 <https://doi.org/10.1080/00263206.2013.850076>
- 4 de Deckker, P., Abed, R. M. M., de Beer, D., Hinrichs, K. U., O’Loingsigh, T., Schefuß, E., ... Van
5 Der Kaars, S. (2008). Geochemical and microbiological fingerprinting of airborne dust that fell
6 in Canberra, Australia, in October 2002. *Geochemistry, Geophysics, Geosystems*, 9(12).
7 <https://doi.org/10.1029/2008GC002091>
- 8 de Jong, R., de Bruin, S., de Wit, A., Schaepman, M. E., & Dent, D. L. (2011). Analysis of monotonic
9 greening and browning trends from global NDVI time-series. *Remote Sensing of Environment*,
10 115(2), 692–702. <https://doi.org/10.1016/J.RSE.2010.10.011>
- 11 de Jong, R., Verbesselt, J., Zeileis, A., & Schaepman, M. (2013). Shifts in Global Vegetation Activity
12 Trends. *Remote Sensing*, 5(3), 1117–1133. <https://doi.org/10.3390/rs5031117>
- 13 de la Paix, M. J., Lanhai, L., Xi, C., Ahmed, S., & Vareniam, A. (2011). Soil Degradation and
14 Altered Flood Risk as a Consequence of Deforestation. *Land Degradation & Development*,
15 24(5), n/a-n/a. <https://doi.org/10.1002/ldr.1147>
- 16 de la Sota, C., Lumbreras, J., Pérez, N., Ealo, M., Kane, M., Youm, I., & Viana, M. (2018). Indoor air
17 pollution from biomass cookstoves in rural Senegal. *Energy for Sustainable Development*, 43,
18 224–234. <https://doi.org/10.1016/J.ESD.2018.02.002>
- 19 de Pauw, E., Saba, M., & Ali, S. (2015). *Mapping climate change in Iraq and Jordan* (ICARDA
20 Working Paper No. 27). Beirut, Lebanon. <https://doi.org/10.13140/RG.2.1.1713.1365>
- 21 de Sherbinin, A., Levy, M., Adamo, S., MacManus, K., Yetman, G., Mara, V., ... Pistolesi, L. (2012).
22 Migration and risk: net migration in marginal ecosystems and hazardous areas. *Environmental*
23 *Research Letters*, 7(4), 045602. <https://doi.org/10.1088/1748-9326/7/4/045602>
- 24 de Vries, F. T., Liiri, M. E., Bjørnlund, L., Bowker, M., Christensen, S., Setälä, H., & Bardgett, R. D.
25 (2012). *Land use alters the resistance and resilience of soil food webs to drought*. *Nature*
26 *Climate Change* (Vol. 2). <https://doi.org/10.1038/nclimate1368>
- 27 Decker, M., Pitman, A. J., Evans, J. P., Decker, M., Pitman, A. J., & Evans, J. P. (2013). Groundwater
28 Constraints on Simulated Transpiration Variability over Southeastern Australian Forests.
29 *Journal of Hydrometeorology*, 14(2), 543–559. <https://doi.org/10.1175/JHM-D-12-058.1>
- 30 Deininger, K., & Jin, S. (2006). Tenure security and land-related investment: Evidence from Ethiopia.
31 *European Economic Review*, 50(5), 1245–1277.
32 <https://doi.org/10.1016/J.EUROECOREV.2005.02.001>
- 33 del Barrio, G., Sanjuan, M., Hirche, A., Yassin, M., Ruiz, A., Ouessar, M., ... Puigdefabregas, J.
34 (2016). Land Degradation States and Trends in the Northwestern Maghreb Drylands, 1998–
35 2008. *Remote Sensing*, 8(7), 603. <https://doi.org/10.3390/rs8070603>
- 36 Dell’Angelo, J., D’Odorico, P., Rulli, M. C., & Marchand, P. (2017). The Tragedy of the Grabbed
37 Commons: Coercion and Dispossession in the Global Land Rush. *World Development*, 92, 1–12.
38 <https://doi.org/10.1016/J.WORLDDEV.2016.11.005>
- 39 Derpsch, R., Friedrich, T., Kassam, A., & Hongwen, L. (2010). Current status of adoption of no-till
40 farming in the world and some of its main benefits. *International Journal of Agricultural and*
41 *Biological Engineering*, 3(1), 1–25. <https://doi.org/10.3965/j.issn.1934-6344.2010.01.0-0>
- 42 Descheemaeker, K., Zijlstra, M., Masikati, P., Crespo, O., & Homann-Kee Tui, S. (2018). Effects of
43 climate change and adaptation on the livestock component of mixed farming systems: A
44 modelling study from semi-arid Zimbabwe. *Agricultural Systems*, 159, 282–295.
45 <https://doi.org/10.1016/J.AGSY.2017.05.004>
- 46 Deshpande, R. S. (2008). Watersheds: Putting the cart before the horse. *Economic and Political*
47 *Weekly*, 74–76.

- 1 Devine, A. P., McDonald, R. A., Quaipe, T., & Maclean, I. M. D. (2017). Determinants of woody
2 encroachment and cover in African savannas. *Oecologia*, 183(4), 939–951.
3 <https://doi.org/10.1007/s00442-017-3807-6>
- 4 Dey, A., Gupta, A., & Singh, G. (2017). Open Innovation at Different Levels for Higher Climate Risk
5 Resilience. *Science, Technology and Society*, 22(3), 388–406.
6 <https://doi.org/10.1177/0971721817723242>
- 7 Dharumarajan, S., Bishop, T. F. A., Hegde, R., & Singh, S. K. (2018). Desertification vulnerability
8 index-an effective approach to assess desertification processes: A case study in Anantapur
9 District, Andhra Pradesh, India. *Land Degradation & Development*, 29(1), 150–161.
10 <https://doi.org/10.1002/ldr.2850>
- 11 Diao, X., & Sarpong, D. B. (2011). Poverty Implications of Agricultural Land Degradation in Ghana:
12 An Economy-wide, Multimarket Model Assessment. *African Development Review*, 23(3), 263–
13 275. <https://doi.org/10.1111/j.1467-8268.2011.00285.x>
- 14 Díaz, J., Linares, C., Carmona, R., Russo, A., Ortiz, C., Salvador, P., & Trigo, R. M. (2017). Saharan
15 dust intrusions in Spain: Health impacts and associated synoptic conditions. *Environmental*
16 *Research*, 156, 455–467. <https://doi.org/10.1016/J.ENVRES.2017.03.047>
- 17 Dibari, C., Bindi, M., Moriondo, M., Staglianò, N., Targetti, S., & Argenti, G. (2016). Spatial data
18 integration for the environmental characterization of pasture macrotypes in the Italian Alps.
19 *Grass and Forage Science*, 71(2), 219–234. <https://doi.org/10.1111/gfs.12168>
- 20 Dile, Y. T., Karlberg, L., Temesgen, M., & Rockström, J. (2013). The role of water harvesting to
21 achieve sustainable agricultural intensification and resilience against water related shocks in
22 sub-Saharan Africa. *Agriculture, Ecosystems and Environment*, 181(2013), 69–79.
23 <https://doi.org/10.1016/j.agee.2013.09.014>
- 24 Dimelu, M., Salifu, E., & Igbokwe, E. (2016). Resource use conflict in agrarian communities,
25 management and challenges: A case of farmer-herdsmen conflict in Kogi State, Nigeria. *Journal*
26 *of Rural Studies*, 46, 147–154. Retrieved from
27 <http://www.sciencedirect.com/science/article/pii/S0743016716300973>
- 28 Dimobe, K., Ouédraogo, A., Soma, S., Goetze, D., Porembski, S., & Thiombiano, A. (2015).
29 Identification of driving factors of land degradation and deforestation in the Wildlife Reserve of
30 Bontioli (Burkina Faso, West Africa). *Global Ecology and Conservation*, 4, 559–571.
31 <https://doi.org/10.1016/j.gecco.2015.10.006>
- 32 Divinsky, I., Becker, N., & Bar (Kutieli), P. (2017). Ecosystem service tradeoff between grazing
33 intensity and other services - A case study in Karei-Deshe experimental cattle range in northern
34 Israel. *Ecosystem Services*, 24(November), 16–27. <https://doi.org/10.1016/j.ecoser.2017.01.002>
- 35 Dixon, J. A., James, D. E., Sherman, P. B., James, D. E., & Sherman, P. B. (2013). *Economics of*
36 *Dryland Management*. Routledge. <https://doi.org/10.4324/9781315066257>
- 37 Djanibekov, U., & Khamzina, A. (2016). Stochastic Economic Assessment of Afforestation on
38 Marginal Land in Irrigated Farming System. *Environmental and Resource Economics*, 63(1),
39 95–117. <https://doi.org/10.1007/s10640-014-9843-3>
- 40 Djanibekov, U., Villamor, G., Dzhakypbekova, K., Chamberlain, J., Xu, J., Djanibekov, U., ... Xu, J.
41 (2016). Adoption of Sustainable Land Uses in Post-Soviet Central Asia: The Case for
42 Agroforestry. *Sustainability*, 8(10), 1030. <https://doi.org/10.3390/su8101030>
- 43 Do, N., & Kang, S. (2014). Assessing drought vulnerability using soil moisture-based water use
44 efficiency measurements obtained from multi-sensor satellite data in Northeast Asia dryland
45 regions. *Journal of Arid Environments*, 105, 22–32.
46 <https://doi.org/10.1016/J.JARIDENV.2014.02.018>
- 47 Dodd, J. (1994). Desertification and Degradation of Africa's Rangelands. *Rangelands*, 16(5).
- 48 Dominguez, F., Cañon, J., & Valdes, J. (2010). IPCC-AR4 climate simulations for the Southwestern

- 1 US: The importance of future ENSO projections. *Climatic Change*, 99(3), 499–514.
2 <https://doi.org/10.1007/s10584-009-9672-5>
- 3 Dominguez, P. (2014). Current situation and future patrimonializing perspectives for the governance
4 of agro-pastoral resources in the Ait Ikis transhumants of the High Atlas (Morocco), 148–166.
5 <https://doi.org/10.4324/9781315768014-14>
- 6 Donat, M. G., Lowry, A. L., Alexander, L. V., O’Gorman, P. A., & Maher, N. (2016). More extreme
7 precipitation in the world’s dry and wet regions. *Nature Climate Change*, 6(5), 508–513.
8 <https://doi.org/10.1038/nclimate2941>
- 9 Dong, S. (2016). Overview: Pastoralism in the World. In B. R. Dong S., Kassam KA., Tourrand J.
10 (Ed.), *Building Resilience of Human-Natural Systems of Pastoralism in the Developing World*
11 (pp. 1–37). Cham: Springer International Publishing. [https://doi.org/doi.org/10.1007/978-3-319-](https://doi.org/doi.org/10.1007/978-3-319-30732-9_1)
12 [30732-9_1](https://doi.org/doi.org/10.1007/978-3-319-30732-9_1)
- 13 Donohue, R. J., McVicar, T. R., & Roderick, M. L. (2009). Climate-related trends in Australian
14 vegetation cover as inferred from satellite observations, 1981–2006. *Global Change Biology*,
15 *15*(4), 1025–1039. <https://doi.org/10.1111/j.1365-2486.2008.01746.x>
- 16 Donohue, R. J., Roderick, M. L., McVicar, T. R., & Farquhar, G. D. (2013). Impact of CO₂
17 fertilization on maximum foliage cover across the globe’s warm, arid environments.
18 *Geophysical Research Letters*, *40*(12), 3031–3035. <https://doi.org/10.1002/grl.50563>
- 19 Dougill, A. J., Akanyang, L., Perkins, J. S., Eckardt, F. D., Stringer, L. C., Favretto, N., ... Mulale, K.
20 (2016). Land use, rangeland degradation and ecological changes in the southern Kalahari,
21 Botswana. *African Journal of Ecology*, *54*(1), 59–67. <https://doi.org/10.1111/aje.12265>
- 22 Dougill, A. J., Fraser, E. D. G., & Reed, M. S. (2010). Anticipating Vulnerability to Climate Change
23 in Dryland Pastoral Systems : Using Dynamic Systems Models for the Kalahari, *15*(2).
- 24 Dow, K., Berkhout, F., Preston, B. L., Klein, R. J. T., Midgley, G., & Shaw, M. R. (2013). Limits to
25 adaptation. *Nature Climate Change*, *3*(4), 305–307. <https://doi.org/10.1038/nclimate1847>
- 26 Dregne, H. E., & Chou, N.-T. (1992). Global Desertification Dimensions and Costs. Retrieved May
27 24, 2018, from <http://www.ciesin.org/docs/002-186/002-186.html>
- 28 Dressler, W., Büscher, B., Schoon, M., Brockington, D., Hayes, T., Kull, C. A., ... Shrestha, K.
29 (2010). From hope to crisis and back again? A critical history of the global CBNRM narrative.
30 *Environmental Conservation*, *37*(01), 5–15. <https://doi.org/10.1017/S0376892910000044>
- 31 Du, F. (2012). Ecological Resettlement of Tibetan Herders in the Sanjiangyuan: A Case Study in
32 Madoi County of Qinghai. *Nomadic Peoples*, *16*(1), 116–133.
33 <https://doi.org/10.3167/np.2012.160109>
- 34 Dukes, J. S., Chiariello, N. R., Loarie, S. R., & Field, C. B. (2011). Strong response of an invasive
35 plant species (*Centaurea solstitialis* L.) to global environmental changes. *Ecological*
36 *Applications*, *21*(6), 1887–1894. <https://doi.org/10.1890/11-0111.1>
- 37 Dukhovny, V. A., Sokolov, V. I., & Ziganshina, D. R. (2013). Integrated Water Resources
38 Management in Central Asia, as a way of survival in conditions of water scarcity. *Quaternary*
39 *International*, *311*, 181–188. <https://doi.org/10.1016/J.QUAINT.2013.07.003>
- 40 Dumenu, W. K., & Obeng, E. A. (2016). Climate change and rural communities in Ghana: Social
41 vulnerability, impacts, adaptations and policy implications. *Environmental Science & Policy*, *55*,
42 208–217. <https://doi.org/10.1016/J.ENVSCI.2015.10.010>
- 43 Durant, S. M., Wachter, T., Bashir, S., Woodroffe, R., De Ornellas, P., Ransom, C., ... Petteorelli, N.
44 (2014). Fiddling in biodiversity hotspots while deserts burn? Collapse of the Sahara’s
45 megafauna. *Diversity and Distributions*, *20*(1), 114–122. <https://doi.org/10.1111/ddi.12157>
- 46 Eekhout, J., & de Vente, J. (2019). Assessing the effectiveness of Sustainable Land Management for
47 large-scale climate change adaptation. *Science of The Total Environment*, *654*, 85–93.

- 1 <https://doi.org/10.1016/J.SCITOTENV.2018.10.350>
- 2 Ehrenfeld, J. G. (2003). Effects of Exotic Plant Invasions on Soil Nutrient Cycling Processes.
3 *Ecosystems*, 6(6), 503–523. <https://doi.org/10.1007/s10021-002-0151-3>
- 4 Eitelberg, D. A., van Vliet, J., & Verburg, P. H. (2015). A review of global potentially available
5 cropland estimates and their consequences for model-based assessments. *Global Change*
6 *Biology*, 21(3), 1236–1248. <https://doi.org/10.1111/gcb.12733>
- 7 El-Fadel, M., Chedid, R., Zeinati, M., & Hmaidan, W. (2003). Mitigating energy-related GHG
8 emissions through renewable energy. *Renewable Energy*, 28(8), 1257–1276.
9 [https://doi.org/10.1016/S0960-1481\(02\)00229-X](https://doi.org/10.1016/S0960-1481(02)00229-X)
- 10 Eldridge, D. J., Bowker, M. A., Maestre, F. T., Roger, E., Reynolds, J. F., & Whitford, W. G. (2011).
11 Impacts of shrub encroachment on ecosystem structure and functioning: Towards a global
12 synthesis. *Ecology Letters*, 14(7), 709–722. <https://doi.org/10.1111/j.1461-0248.2011.01630.x>
- 13 Eldridge, D. J., & Soliveres, S. (2014). Are shrubs really a sign of declining ecosystem function?
14 Disentangling the myths and truths of woody encroachment in Australia. *Australian Journal of*
15 *Botany*, 62(7), 594–608. <https://doi.org/10.1071/BT14137>
- 16 Eldridge, D. J., Wang, L., & Ruiz-Colmenero, M. (2015). Shrub encroachment alters the spatial
17 patterns of infiltration. *Ecohydrology*, 8(1), 83–93. <https://doi.org/10.1002/eco.1490>
- 18 Elhadary, Y. (2014). Examining drivers and indicators of the recent changes among pastoral
19 communities of butana locality, gedarif State, Sudan. *American Journal of Sociological*
20 *Research*, 4(3), 88–101. Retrieved from
21 <http://article.sapub.org/10.5923.j.sociology.20140403.04.html>
- 22 Ellies, A. (2000). Soil erosion and its control in Chile. *Acta Geológica Hispana*, 35, 279–284.
23 Retrieved from <http://www.raco.cat/>
- 24 Ellis, J. (1995). Climate variability and complex ecosystem dynamics: implications for pastoral
25 development. In *Living with Uncertainty* (pp. 37–46). Rugby, Warwickshire, United Kingdom:
26 Practical Action Publishing. <https://doi.org/10.3362/9781780445335.002>
- 27 Elmore, A. J., Kaste, J. M., Okin, G. S., & Fantle, M. S. (2008). Groundwater influences on
28 atmospheric dust generation in deserts. *Journal of Arid Environments*, 72(10), 1753–1765.
29 <https://doi.org/10.1016/J.JARIDENV.2008.05.008>
- 30 Elum, Z., & AS Momodu. (2017). Climate change mitigation and renewable energy for sustainable
31 development in Nigeria: A discourse approach. *Renewable and Sustainable Energy Reviews*, 76,
32 72–80. Retrieved from <http://www.sciencedirect.com/science/article/pii/S1364032117303544>
- 33 Enfors, E. I., & Gordon, L. J. (2008). Dealing with drought: The challenge of using water system
34 technologies to break dryland poverty traps. *Global Environmental Change*, 18(4), 607–616.
35 <https://doi.org/10.1016/J.GLOENVCHA.2008.07.006>
- 36 Engdawork, A., & Bork, H.-R. (2016). Farmers' Perception of Land Degradation and Traditional
37 Knowledge in Southern Ethiopia-Resilience and Stability. *Land Degradation & Development*,
38 27(6), 1552–1561. <https://doi.org/10.1002/ldr.2364>
- 39 Engelbrecht, F. A., Landman, W. A., Engelbrecht, C. J., Landman, S., Bopape, M. M., Roux, B., ...
40 Thatcher, M. (2011). Multi-scale climate modelling over Southern Africa using a variable-
41 resolution global model. In *Water Research Commission 40-year Celebration Conference* (Vol.
42 37, pp. 647–658). Kempton Park.
- 43 Engelstaedter, S., Kohfeld, K. E., Tegen, I., & Harrison, S. P. (2003). Controls of dust emissions by
44 vegetation and topographic depressions: An evaluation using dust storm frequency data.
45 *Geophysical Research Letters*, 30(6). <https://doi.org/10.1029/2002GL016471>
- 46 Erb, K.-H., Fetzel, T., Kastner, T., Kroisleitner, C., Lauk, C., Mayer, A., & Niedertscheider, M.
47 (2016). Livestock Grazing, the Neglected Land Use. In *Social Ecology* (pp. 295–313). Cham:

- 1 Springer International Publishing. https://doi.org/10.1007/978-3-319-33326-7_13
- 2 Eriksen, S., & Lind, J. (2009). Adaptation as a Political Process: Adjusting to Drought and Conflict in
3 Kenya's Drylands. *Environmental Management*, 43(5), 817–835.
4 <https://doi.org/10.1007/s00267-008-9189-0>
- 5 Etana, A., Kelbessa, E., & Soromessa, T. (2011). Impact of Parthenium hysterophorus L.(Asteraceae)
6 on herbaceous plant biodiversity of Awash National Park (ANP), Ethiopia. *Management of*
7 *Biological Invasions*, 2. <https://doi.org/10.3391/mbi.2011.2.1.07>
- 8 Etchart, L. (2017). The role of indigenous peoples in combating climate change. *Palgrave*
9 *Communications*, 3, 17085. <https://doi.org/10.1057/palcomms.2017.85>
- 10 Etongo, D., Djenontin, I., Kanninen, M., & Fobissie, K. (2015). Smallholders' Tree Planting Activity
11 in the Ziro Province, Southern Burkina Faso: Impacts on Livelihood and Policy Implications.
12 *Forests*, 6(12), 2655–2677. <https://doi.org/10.3390/f6082655>
- 13 European Court of Auditors. (2018). *Desertification in Europe. Background paper*.
- 14 European Court of Auditors (ECA). (2018). *Combating desertification in the EU: a growing threat in*
15 *need of more action*. Retrieved from
16 https://www.eca.europa.eu/Lists/ECADocuments/SR18_33/SR_DESERTIFICATION_EN.pdf
- 17 Evan, A. T. (2018). Surface Winds and Dust Biases in Climate Models. *Geophysical Research*
18 *Letters*, 45(2), 1079–1085. <https://doi.org/10.1002/2017GL076353>
- 19 Evan, A. T., Flamant, C., Fiedler, S., & Doherty, O. (2014). An analysis of aeolian dust in climate
20 models. *Geophysical Research Letters*, 41(16), 5996–6001.
21 [https://doi.org/10.1002/2014GL060545@10.1002/\(ISSN\)1944-8007.GRLCMIP5](https://doi.org/10.1002/2014GL060545@10.1002/(ISSN)1944-8007.GRLCMIP5)
- 22 Evan, A. T., Flamant, C., Gaetani, M., & Guichard, F. (2016). The past, present and future of African
23 dust. *Nature*, 531(7595), 493–495. <https://doi.org/10.1038/nature17149>
- 24 Evan, A. T., & Mukhopadhyay, S. (2010). African Dust over the Northern Tropical Atlantic: 1955–
25 2008. *Journal of Applied Meteorology and Climatology*, 49(11), 2213–2229.
26 <https://doi.org/10.1175/2010JAMC2485.1>
- 27 Evan, A. T., Vimont, D. J., Heidinger, A. K., Kossin, J. P., & Bennartz, R. (2009). The role of
28 aerosols in the evolution of tropical North Atlantic Ocean temperature anomalies. *Science (New*
29 *York, N.Y.)*, 324(5928), 778–81. <https://doi.org/10.1126/science.1167404>
- 30 Evans, J., & Geerken, R. (2004). Discrimination between climate and human-induced dryland
31 degradation. *Journal of Arid Environments*, 57(4), 535–554. [https://doi.org/10.1016/S0140-1963\(03\)00121-6](https://doi.org/10.1016/S0140-1963(03)00121-6)
- 32
- 33 Evans, J. P., & Geerken, R. (2006). Classifying rangeland vegetation type and coverage using a
34 Fourier component based similarity measure. *Remote Sensing of Environment*, 105(1), 1–8.
35 <https://doi.org/10.1016/j.rse.2006.05.017>
- 36 Evans, J. P., Meng, X., & McCabe, M. F. (2017). Land surface albedo and vegetation feedbacks
37 enhanced the millennium drought in south-east Australia. *Hydrology and Earth System Sciences*,
38 21(1), 409–422. <https://doi.org/10.5194/hess-21-409-2017>
- 39 Evans, L. S., Kiem, A. S., Pickering, C. M., Barnett, J., Palutikof, J. P., Kingsford, R. T., ... Smithers,
40 S. G. (2015). From barriers to limits to climate change adaptation: path dependency and the
41 speed of change. *Ecology and Society*, 20(3). <https://doi.org/10.5751/es-07698-200305>
- 42 Ezra, M., & Kiros, G.-E. (2001). Rural Out-migration in the Drought Prone Areas of Ethiopia: A
43 Multilevel Analysis1. *International Migration Review*, 35(3), 749–771.
44 <https://doi.org/10.1111/j.1747-7379.2001.tb00039.x>
- 45 Fan, B., Guo, L., Li, N., Chen, J., Lin, H., Zhang, X., ... Ma, L. (2015). Earlier vegetation green-up
46 has reduced spring dust storms. *Scientific Reports*, 4(1), 6749. <https://doi.org/10.1038/srep06749>

- 1 FAO. (1995). *Desertification and drought - extent and consequences proposal for a participatory*
2 *approach to combat desertification*. Rome. Retrieved from
3 <http://www.fao.org/docrep/x5317e/x5317e00.htm#Contents>
- 4 FAO. (2004). *Drought impact mitigation and prevention in the Limpopo River Basin. Drought impact*
5 *mitigation and prevention in the Limpopo River Basin A situation analysis*. Rome.
- 6 FAO. (2011). *Gestion des plantations sur dunes. Document de travail sur les Forêts et la Foresterie*
7 *en zones arides*. Retrieved from www.fao.org/forestry/aridzone
- 8 FAO. (2015). *Global soil status, processes and trends. Status of the World's Soil Resources*.
- 9 FAO. (2016). *Trees, forests and land use in drylands The first global assessment Preliminary findings*
10 (First). Rome: FAO. Retrieved from <http://www.fao.org/3/a-i5905e.pdf>
- 11 FAO. (2018). *The impact of disasters and crises on agriculture and food security 2017*. Rome, Italy.
12 Retrieved from <http://www.fao.org/3/I8656EN/i8656en.pdf>
- 13 FAO, IFAD, UNICEF, WFP, & WHO. (2018). *The State of Food Security and Nutrition in the World*
14 *2018. Building climate resilience for food security and nutrition*. Rome, Italy. Retrieved from
15 <http://www.fao.org/3/I9553EN/i9553en.pdf>
- 16 Faour-Klingbeil, D., & Todd, E. C. D. (2018). The Impact of Climate Change on Raw and Untreated
17 Wastewater Use for Agriculture, Especially in Arid Regions: A Review. *Foodborne Pathogens*
18 *and Disease*, 15(2), 61–72. <https://doi.org/10.1089/fpd.2017.2389>
- 19 Farage, P. K., Ardo, J., Olsson, L., Rienzi, E. A., Ball, A. S., & Pretty, J. N. (2007). The potential for
20 soil carbon sequestration in three tropical dryland farming systems of Africa and Latin America:
21 A modelling approach. *Soil and Tillage Research*, (94), 457–472.
- 22 Farnoosh, A., Lantz, F., & Percebois, J. (2014). Electricity generation analyses in an oil-exporting
23 country: Transition to non-fossil fuel based power units in Saudi Arabia. *Energy*, 69, 299–308.
24 <https://doi.org/10.1016/J.ENERGY.2014.03.017>
- 25 Fasil, R. (2011). Parthenium hysterophorus in Ethiopia : Distribution and Importance, and Current
26 Efforts to Manage the Scourge. Workshop on noxious weeds in production of certified seeds,
27 11-12 July 2011, FAO, Accra, Ghana.
- 28 Favet, J., Lapanje, A., Giongo, A., Kennedy, S., Aung, Y. Y., Cattaneo, A., ... Gorbushina, A. A.
29 (2013). Microbial hitchhikers on intercontinental dust: Catching a lift in Chad. *ISME Journal*,
30 7(4), 850–867. <https://doi.org/10.1038/ismej.2012.152>
- 31 Fay, P. A., Kaufman, D. M., Nippert, J. B., Carlisle, J. D., & Harper, C. W. (2008). Changes in
32 grassland ecosystem function due to extreme rainfall events: Implications for responses to
33 climate change. *Global Change Biology*, 14(7), 1600–1608. [https://doi.org/10.1111/j.1365-](https://doi.org/10.1111/j.1365-2486.2008.01605.x)
34 [2486.2008.01605.x](https://doi.org/10.1111/j.1365-2486.2008.01605.x)
- 35 Feng, S., Krueger, A. B., & Oppenheimer, M. (2010). Linkages among climate change, crop yields
36 and Mexico–US cross-border migration. *National Acad Sciences*. Retrieved from
37 <http://www.pnas.org/content/107/32/14257.short>
- 38 Feng, Q., Ma, H., Jiang, X., Wang, X., & Cao, S. (2015). What Has Caused Desertification in China?
39 *Scientific Reports*, 5, 15998. <https://doi.org/10.1038/srep15998>
- 40 Feng, S., & Fu, Q. (2013). Expansion of global drylands under a warming climate. *Atmospheric*
41 *Chemistry and Physics*, 13(19), 10081–10094. <https://doi.org/10.5194/acp-13-10081-2013>
- 42 Feng, S., Hu, Q., Huang, W., Ho, C.-H., Li, R., & Tang, Z. (2014). Projected climate regime shift
43 under future global warming from multi-model, multi-scenario CMIP5 simulations. *Global and*
44 *Planetary Change*, 112, 41–52. <https://doi.org/10.1016/J.GLOPLACHA.2013.11.002>
- 45 Feng, X., Fu, B., Piao, S., Wang, S., Ciais, P., Zeng, Z., ... Wu, B. (2016). Revegetation in China's
46 Loess Plateau is approaching sustainable water resource limits. *Nature Climate Change*, 6(11),
47 1019–1022. <https://doi.org/10.1038/nclimate3092>

- 1 Fensholt, R., Horion, S., Tagesson, T., Ehammer, A., Grogan, K., Tian, F., ... Rasmussen, K. (2015).
2 Assessment of Vegetation Trends in Drylands from Time Series of Earth Observation Data (pp.
3 159–182). Springer, Cham. https://doi.org/10.1007/978-3-319-15967-6_8
- 4 Fensholt, R., Langanke, T., Rasmussen, K., Reenberg, A., Prince, S. D., Tucker, C., ... Wessels, K.
5 (2012). Greenness in semi-arid areas across the globe 1981–2007 — an Earth Observing
6 Satellite based analysis of trends and drivers. *Remote Sensing of Environment*, *121*, 144–158.
7 <https://doi.org/10.1016/J.RSE.2012.01.017>
- 8 Fernandes, P. M., Davies, G. M., Ascoli, D., Fernández, C., Moreira, F., Rigolot, E., ... Molina, D.
9 (2013). Prescribed burning in southern Europe: developing fire management in a dynamic
10 landscape. *Frontiers in Ecology and the Environment*, *11*(s1), e4–e14.
11 <https://doi.org/10.1890/120298>
- 12 Fernandez-Gimenez, M., Batkhishig, B., & B Batbuyan. (2015). Lessons from the dzud: Community-
13 based rangeland management increases the adaptive capacity of Mongolian herders to winter
14 disasters. *World Development*, *68*, 48–65. Retrieved from
15 <http://www.sciencedirect.com/science/article/pii/S0305750X14003738>
- 16 Fernandez-Gimenez, M. E. (2000). The Role of Mongolian Nomadic Pastoralists' Ecological
17 Knowledge in Rangeland Management. *Ecological Applications*, *10*(5), 1318–1326.
18 [https://doi.org/10.1890/1051-0761\(2000\)010\[1318:TROMNP\]2.0.CO;2](https://doi.org/10.1890/1051-0761(2000)010[1318:TROMNP]2.0.CO;2)
- 19 Fernández-Giménez, M. E., & Fillat Estaque, F. (2012). Pyrenean Pastoralists' Ecological
20 Knowledge: Documentation and Application to Natural Resource Management and Adaptation.
21 *Human Ecology*, *40*(2), 287–300. <https://doi.org/10.1007/s10745-012-9463-x>
- 22 Fernández, O. A., Gil, M. E., & Distel, R. A. (2009). The challenge of rangeland degradation in a
23 temperate semiarid region of Argentina: The Caldenal. *Land Degradation & Development*,
24 *20*(4), 431–440. <https://doi.org/10.1002/ldr.851>
- 25 Fetzl, T., Havlik, P., Herrero, M., Kaplan, J. O., Kastner, T., Kroisleitner, C., ... Erb, K.-H. (2017).
26 Quantification of uncertainties in global grazing systems assessment. *Global Biogeochemical*
27 *Cycles*, *31*(7), 1089–1102. <https://doi.org/10.1002/2016GB005601>
- 28 Few, R., & Tebboth, M. G. L. (2018). Recognising the dynamics that surround drought impacts.
29 *Journal of Arid Environments*, *157*, 113–115. <https://doi.org/10.1016/J.JARIDENV.2018.06.001>
- 30 Ficklin, D. L., Abatzoglou, J. T., Robeson, S. M., & Dufficy, A. (2016). The influence of climate
31 model biases on projections of aridity and drought. *Journal of Climate*, *29*(4), 1369–1389.
32 <https://doi.org/10.1175/JCLI-D-15-0439.1>
- 33 Field, J. P., Belnap, J., Breshears, D. D., Neff, J. C., Okin, G. S., Whicker, J. J., ... & Reynolds, R. L.
34 (2010, October). The ecology of dust. *Frontiers in Ecology and the Environment*. Wiley-
35 Blackwell. <https://doi.org/10.1890/090050>
- 36 Field, C. B., Barros, V. R., Mach, K. J., Mastrandrea, M. D., Aalst, M. van, Adger, W. N., ... Yohe,
37 G. W. (2014). Technical summary. In V. R. B. Field, C.B., B. G. D.J. Dokken, K.J. Mach, M.D.
38 Mastrandrea, T.E. Bilir, M. Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova, & and L. L. W.
39 E.S. Kissel, A.N. Levy, S. MacCracken, P.R. Mastrandrea (Eds.), *Climate Change 2014:*
40 *Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of*
41 *Working Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate*
42 *Change*. Cambridge University Press. Retrieved from
43 https://www.ipcc.ch/site/assets/uploads/2018/02/WGIIAR5-TS_FINAL.pdf
- 44 Findlay, A. M. (2011). Migrant destinations in an era of environmental change. *Global Environmental*
45 *Change*, *21*, S50–S58. <https://doi.org/10.1016/J.GLOENVCHA.2011.09.004>
- 46 Fischlin, A., Midgley, G. F., Price, J. T., Leemans, R., Gopal, B., Turley, C., ... Velichko, A. A.
47 (2007). Ecosystems, their properties, goods, and services. *Climate Change 2007: Impacts,*
48 *Adaptation and Vulnerability. Contribution of Working Group II to the Fourth Assessment*
49 *Report of the Intergovernmental Panel on Climate Change*. In M. L. Parry, O. F. Canziani, J. P.

- 1 Palutikof, P. J. van der Linden, & C. E. Hanson (Eds.) (pp. 211–272). Cambridge University
2 Press, Cambridge, UK.
- 3 Fleischner, T. L. (1994). Costos ecológicos del pastoreo de ganado en el oeste de Estados Unidos.
4 *Conservation Biology*, 8(3), 629–644. <https://doi.org/10.1046/j.1523-1739.1994.08030629.x>
- 5 Fleskens, L., & Stringer, L. C. (2014). Land Management and Policy Responses to Mitigate
6 Desertification and Land Degradation. *Land Degradation & Development*, 25(1), 1–4.
7 <https://doi.org/10.1002/ldr.2272>
- 8 Fleurbaey, M., Kartha, S., Bolwig, S., Chee, Y. L., Chen, Y., Corbera, E., ... Sagar, A. D. (2014).
9 Sustainable Development and Equity. In O. Edenhofer, R. Pichs-Madruga, Y. Sokona, E.
10 Farahani, S. Kadner, K. Seyboth, ... J. C. Minx (Eds.), *Climate Change 2014: Mitigation of*
11 *Climate Change. Contribution of Working Group III to the Fifth Assessment Report of the*
12 *Intergovernmental Panel on Climate Change* (pp. 238–350). Cambridge, United Kingdom and
13 New York, NY, USA: Cambridge University Press.
- 14 Ford, T. W., & Labosier, C. F. (2017). Meteorological conditions associated with the onset of flash
15 drought in the Eastern United States. *Agricultural and Forest Meteorology*, 247, 414–423.
16 <https://doi.org/10.1016/J.AGRFORMET.2017.08.031>
- 17 Franklin, K., & Molina-Freaner, F. (2010). Consequences of buffelgrass pasture development for
18 primary productivity, perennial plant richness, and vegetation structure in the drylands of
19 sonora, Mexico. *Conservation Biology*, 24(6), 1664–1673. [https://doi.org/10.1111/j.1523-](https://doi.org/10.1111/j.1523-1739.2010.01540.x)
20 [1739.2010.01540.x](https://doi.org/10.1111/j.1523-1739.2010.01540.x)
- 21 Franks, P. J., Adams, M. A., Amthor, J. S., Barbour, M. M., Berry, J. A., Ellsworth, D. S., ... von
22 Caemmerer, S. (2013). Sensitivity of plants to changing atmospheric CO₂ concentration: from
23 the geological past to the next century. *New Phytologist*, 197(4), 1077–1094.
24 <https://doi.org/10.1111/nph.12104>
- 25 Fraser, E., Dougill, A., K Hubacek, Quinn, C., Sendzimir, J., & Termansen, M. (2011). Assessing
26 vulnerability to climate change in dryland livelihood systems: conceptual challenges and
27 interdisciplinary solutions. *Ecology and Society*, 16(3). Retrieved from
28 <https://www.ecologyandsociety.org/vol16/iss3/art3/main.html>
- 29 Fratkin, E. (2013). Seeking alternative livelihoods in pastoral areas. In Andy Catley, Jeremy Lind, &
30 Ian Scoones (Eds.), *Pastoralism and Development in Africa Dynamic Change at the Margins*
31 (First). Routledge.
- 32 Fre, Z. (2018). *Knowledge Sovereignty among African Cattle Herders*. London: UCL Press.
33 <https://doi.org/https://doi.org/10.14324/111.9781787353114>
- 34 Fried, T., Weissbrod, L., Tepper, Y., & Bar-Oz, G. (2018). A glimpse of an ancient agricultural
35 ecosystem based on remains of micromammals in the Byzantine Negev Desert. *Royal Society*
36 *Open Science*, 5(1), 171528. <https://doi.org/10.1098/rsos.171528>
- 37 Fu, B., Chen, L., Ma, K., Zhou, H., & Wang, J. (2000). The relationships between land use and soil
38 conditions in the hilly area of the loess plateau in northern Shaanxi, China. *CATENA*, 39(1), 69–
39 78. [https://doi.org/10.1016/S0341-8162\(99\)00084-3](https://doi.org/10.1016/S0341-8162(99)00084-3)
- 40 Fu, Q., Johanson, C. M., Wallace, J. M., & Reichler, T. (2006). Enhanced mid-latitude tropospheric
41 warming in satellite measurements. *Science (New York, N.Y.)*, 312(5777), 1179.
42 <https://doi.org/10.1126/science.1125566>
- 43 Fu, Q., Li, B., Hou, Y., Bi, X., & Zhang, X. (2017). Effects of land use and climate change on
44 ecosystem services in Central Asia's arid regions: A case study in Altay Prefecture, China.
45 *Science of The Total Environment*, 607–608, 633–646.
46 <https://doi.org/10.1016/J.SCITOTENV.2017.06.241>
- 47 Fu, Q., Lin, L., Huang, J., Feng, S., & Gettelman, A. (2016). Changes in terrestrial aridity for the
48 period 850–2080 from the community earth system model. *Journal of Geophysical Research*,

- 1 121(6), 2857–2873. <https://doi.org/10.1002/2015JD024075>
- 2 Fuentealba, E., Ferrada, P., Araya, F., Marzo, A., Parrado, C., & Portillo, C. (2015). Photovoltaic
3 performance and LCoE comparison at the coastal zone of the Atacama Desert, Chile. *Energy*
4 *Conversion and Management*, 95, 181–186.
5 <https://doi.org/10.1016/J.ENCONMAN.2015.02.036>
- 6 Fuller, A., Mitchell, D., Maloney, S. K., & Hetem, R. S. (2016). Towards a mechanistic understanding
7 of the responses of large terrestrial mammals to heat and aridity associated with climate change.
8 *Climate Change Responses*, 3(1), 10. <https://doi.org/10.1186/s40665-016-0024-1>
- 9 Funk, C., Harrison, L., Shukla, S., Pomposi, C., Galu, G., Korecha, D., ... Verdin, J. (2018).
10 Examining the role of unusually warm Indo-Pacific sea-surface temperatures in recent African
11 droughts. *Quarterly Journal of the Royal Meteorological Society*, 144(June 2017), 360–383.
12 <https://doi.org/10.1002/qj.3266>
- 13 Funk, C., Pedreros, D., Nicholson, S., Hoell, A., Korecha, D., Galu, G., ... Pomposi, C. (2019).
14 Examining the Potential Contributions of Extreme “Western V” Sea Surface Temperatures to the
15 2017 March–June East African Drought. *Bulletin of the American Meteorological Society*,
16 100(1), S55–S60. <https://doi.org/10.1175/BAMS-D-18-0108.1>
- 17 Gad, M.I.M. and Abdel-Baki, A. A. (2002). Estimation of salt balance in the soil-water of the old
18 cultivated lands in Siwa oasis, western desert, Egypt. In *Proceedings of the International*
19 *Symposium on Optimum Resources Utilization in Salt-Affected Ecosystems in Arid and Semi-*
20 *Arid Regions* (pp. 6–23).
- 21 Galvez, A., Sinicco, T., Cayuela, M. L., Mingorance, M. D., Fornasier, F., & Mondini, C. (2012).
22 Short term effects of bioenergy by-products on soil C and N dynamics, nutrient availability and
23 biochemical properties. *Agriculture, Ecosystems & Environment*, 160, 3–14.
24 <https://doi.org/10.1016/J.AGEE.2011.06.015>
- 25 Gammoh, I. A., & Oweis, T. Y. (2011). *Performance and Adaptation of the Vallerani Mechanized*
26 *Water Harvesting System in Degraded Badia Rangelands*. *Journal of Environmental Science*
27 *and Engineering* (Vol. 5). Retrieved from
28 <http://www.davidpublisher.org/Public/uploads/Contribute/55278f4be0cae.pdf>
- 29 García-Fayos, P., & Bochet, E. (2009). Indication of antagonistic interaction between climate change
30 and erosion on plant species richness and soil properties in semiarid Mediterranean ecosystems.
31 *Global Change Biology*, 15(2), 306–318. <https://doi.org/10.1111/j.1365-2486.2008.01738.x>
- 32 García-Ruiz, J. M. (2010). The effects of land uses on soil erosion in Spain: A review. *Catena*, 81(1),
33 1–11. <https://doi.org/10.1016/j.catena.2010.01.001>
- 34 García-Ruiz, J. M., Beguería, S., Lana-Renault, N., Nadal-Romero, E., & Cerdà, A. (2017). Ongoing
35 and Emerging Questions in Water Erosion Studies. *Land Degradation & Development*, 28(1), 5–
36 21. <https://doi.org/10.1002/ldr.2641>
- 37 García-Ruiz, J. M., Beguería, S., Nadal-Romero, E., González-Hidalgo, J. C., Lana-Renault, N., &
38 Sanjuán, Y. (2015). A meta-analysis of soil erosion rates across the world. *Geomorphology*, 239,
39 160–173. <https://doi.org/10.1016/j.geomorph.2015.03.008>
- 40 Gari, L. (2006). A History of the Hima Conservation System. *Environment and History*, 12(2), 213–
41 228. <https://doi.org/10.3197/096734006776680236>
- 42 Garrity, D. P., Akinnifesi, F. K., Ajayi, O. C., Weldesemayat, S. G., Mowo, J. G., Kalinganire, A., ...
43 Bayala, J. (2010). Evergreen Agriculture: a robust approach to sustainable food security in
44 Africa. *Food Security*, 2(3), 197–214. <https://doi.org/10.1007/s12571-010-0070-7>
- 45 Garschagen, M., & Romero-Lankao, P. (2015). Exploring the relationships between urbanization
46 trends and climate change vulnerability. *Climatic Change*, 133(1), 37–52.
47 <https://doi.org/10.1007/s10584-013-0812-6>
- 48 Gbeckor-Kove, N. (1989). *Drought and Desertification*. Retrieved from

- 1 https://library.wmo.int/pmb_ged/wmo-td_286_en.pdf
- 2 Gbetibouo, G. A., Hassan, R. M., & Ringler, C. (2010). Modelling farmers' adaptation strategies for
3 climate change and variability: The case of the Limpopo Basin, South Africa. *Agrekon*, 49(2),
4 217–234. <https://doi.org/10.1080/03031853.2010.491294>
- 5 Ge, X., Li, Y., Luloff, A. E., Dong, K., & Xiao, J. (2015). Effect of agricultural economic growth on
6 sandy desertification in Horqin Sandy Land. *Ecological Economics*, 119, 53–63.
7 <https://doi.org/10.1016/J.ECOLECON.2015.08.006>
- 8 Gebeye, B. A. (2016). Unsustain the sustainable: An evaluation of the legal and policy interventions
9 for pastoral development in Ethiopia. *Pastoralism*, 6(1). <https://doi.org/10.1186/s13570-016-0049-x>
10
- 11 Gebreselassie, S., Kirui, O. K., & Mirzabaev, A. (2016). Economics of Land Degradation and
12 Improvement in Ethiopia. In *Economics of Land Degradation and Improvement – A Global*
13 *Assessment for Sustainable Development* (pp. 401–430). Cham: Springer International
14 Publishing. https://doi.org/10.1007/978-3-319-19168-3_14
- 15 Geerken, R. A. (2009). An algorithm to classify and monitor seasonal variations in vegetation
16 phenologies and their inter-annual change. *ISPRS Journal of Photogrammetry and Remote*
17 *Sensing*, 64(4), 422–431. <https://doi.org/10.1016/J.ISPRSJPRS.2009.03.001>
- 18 Geerken, R., & Ilaiwi, M. (2004). Assessment of rangeland degradation and development of a strategy
19 for rehabilitation. *Remote Sensing of Environment*, 90(4), 490–504.
20 <https://doi.org/10.1016/J.RSE.2004.01.015>
- 21 Geerken, R., Zaitchik, B., & Evans, J. P. (2005). Classifying rangeland vegetation type and coverage
22 from NDVI time series using Fourier Filtered Cycle Similarity. *International Journal of Remote*
23 *Sensing*, 26(24), 5535–5554. <https://doi.org/10.1080/01431160500300297>
- 24 Geerts, S., & Raes, D. (2009). Deficit irrigation as an on-farm strategy to maximize crop water
25 productivity in dry areas. *Agricultural Water Management*, 96(9), 1275–1284.
26 <https://doi.org/10.1016/J.AGWAT.2009.04.009>
- 27 Geist, H. H. ., & Lambin, E. F. (2004). Dynamic Causal Patterns of Desertification. *BioScience*, 54(9),
28 817–829. [https://doi.org/10.1641/0006-3568\(2004\)054\[0817:DCPOD\]2.0.CO;2](https://doi.org/10.1641/0006-3568(2004)054[0817:DCPOD]2.0.CO;2)
- 29 Gemenne, F. (2011). Why the numbers don't add up: A review of estimates and predictions of people
30 displaced by environmental changes. *Global Environmental Change*, 21, S41–S49.
31 <https://doi.org/10.1016/J.GLOENVCHA.2011.09.005>
- 32 General Directorate of Combating Desertification and Erosion (ÇEMGM). (2017). *Combating*
33 *Desertification and Erosion Activities in Turkey*. Ankara, Turkey. . Retrieved from
34 http://www.cem.gov.tr/erozyon/Libraries/cemDocument/FAAL_ING_1000_AD.sflb.ashx
- 35 Gerber, N., & Mirzabaev, A. (2017). *Benefits of action and costs of inaction: Drought mitigation and*
36 *preparedness – a literature review* (No. 1). WMO, Geneva, Switzerland and GWP, Stockholm,
37 Sweden.
- 38 Gerber, N., Nkonya, E., & von Braun, J. (2014). Land Degradation, Poverty and Marginality. In
39 *Marginality* (pp. 181–202). Dordrecht: Springer Netherlands. https://doi.org/10.1007/978-94-007-7061-4_12
40
- 41 German, L., Mansoor, H., Alemu, G., Mazengia, W., Amede, T., & Stroud, A. (2007). Participatory
42 integrated watershed management: Evolution of concepts and methods in an ecoregional
43 program of the eastern African highlands. *Agricultural Systems*, 94(2), 189–204.
44 <https://doi.org/10.1016/J.AGSY.2006.08.008>
- 45 Ghazi, S., Sayigh, A., & Ip, K. (2014). Dust effect on flat surfaces – A review paper. *Renewable and*
46 *Sustainable Energy Reviews*, 33, 742–751. <https://doi.org/10.1016/J.RSER.2014.02.016>
- 47 Gherboudj, I., Naseema Beegum, S., & Ghedira, H. (2017). Identifying natural dust source regions

- 1 over the Middle-East and North-Africa: Estimation of dust emission potential. *Earth-Science*
2 *Reviews*, 165, 342–355. <https://doi.org/10.1016/J.EARSCIREV.2016.12.010>
- 3 Gholami, A., Alemrajabi, A. A., & Saboonchi, A. (2017). Experimental study of self-cleaning
4 property of titanium dioxide and nanospray coatings in solar applications. *Solar Energy*, 157,
5 559–565. <https://doi.org/10.1016/J.SOLENER.2017.08.075>
- 6 Giannadaki, D., Pozzer, A., & Lelieveld, J. (2014). Modeled global effects of airborne desert dust on
7 air quality and premature mortality. *Atmospheric Chemistry and Physics*, 14(2), 957–968.
8 <https://doi.org/10.5194/acp-14-957-2014>
- 9 Giannakopoulos, C., Le Sager, P., Bindi, M., Moriondo, M., Kostopoulou, E., & Goodess, C. M.
10 (2009). Climatic changes and associated impacts in the Mediterranean resulting from a 2 °C
11 global warming. *Global and Planetary Change*, 68(3), 209–224. Retrieved from
12 <https://www.sciencedirect.com/science/article/pii/S0921818109001131>
- 13 Giannini, A., Biasutti, M., Held, I. M. I. M. I. M., & Sobel, A. H. A. H. A. H. (2008). A global
14 perspective on African climate. *Climatic Change*, 90(4), 359–383.
15 <https://doi.org/10.1007/s10584-008-9396-y>
- 16 Gibbs, H. K., & Salmon, J. M. (2015). Mapping the world ' s degraded lands. *Applied Geography*, 57,
17 12–21. <https://doi.org/10.1016/j.apgeog.2014.11.024>
- 18 Gibson, C. C., Williams, J. T., & Ostrom, E. (2005). Local Enforcement and Better Forests. *World*
19 *Development*, 33(2), 273–284. <https://doi.org/10.1016/J.WORLDDEV.2004.07.013>
- 20 Gichenje, H., & Godinho, S. (2018). Establishing a land degradation neutrality national baseline
21 through trend analysis of GIMMS NDVI Time-series. *Land Degradation & Development*, 29(9),
22 2985–2997. <https://doi.org/10.1002/ldr.3067>
- 23 Giger, M., Liniger, H., Sauter, C., & Schwilch, G. (2015). Economic Benefits and Costs of
24 Sustainable Land Management Technologies: An Analysis of WOCAT's Global Data. *Land*
25 *Degradation & Development*, August 201(4), 962–974. <https://doi.org/10.1002/ldr.2429>
- 26 Ginoux, P., Prospero, J. M., Gill, T. E., Hsu, N. C., & Zhao, M. (2012). Global-scale attribution of
27 anthropogenic and natural dust sources and their emission rates based on MODIS Deep Blue
28 aerosol products. *Reviews of Geophysics*, 50(3). <https://doi.org/10.1029/2012RG000388>
- 29 Gitay, H., Suárez, A., Watson, R. T., & Dokken, D. J. (2002). *Climate change and biodiversity*.
30 Retrieved from <https://www.ipcc.ch/pdf/technical-papers/climate-changes-biodiversity-en.pdf>
- 31 Glendenning, C. J., Van Ogtrop, F. F., Mishra, A. K., & Vervoort, R. W. (2012). Balancing watershed
32 and local scale impacts of rain water harvesting in India—A review. *Agricultural Water*
33 *Management*, 107, 1–13.
- 34 Godfray, H. C. J., & Garnett, T. (2014). Food security and sustainable intensification. *Philosophical*
35 *Transactions of the Royal Society B: Biological Sciences*, 369(1639), 20120273–20120273.
36 <https://doi.org/10.1098/rstb.2012.0273>
- 37 Goffner, D., Sinare, H., & Gordon, L. J. (2019). The Great Green Wall for the Sahara and the Sahel
38 Initiative as an opportunity to enhance resilience in Sahelian landscapes and livelihoods.
39 *Regional Environmental Change*, 1–12. <https://doi.org/10.1007/s10113-019-01481-z>
- 40 Gomes, N. (2006). *Access to water, pastoral resource management and pastoralists' livelihoods:*
41 *Lessons learned from water development in selected areas of Eastern Africa (Kenya, Ethiopia,*
42 *Somalia)* (LSP Working Paper No. 26). Retrieved from [http://www.celep.info/wp-](http://www.celep.info/wp-content/uploads/2012/08/2006-Gomez-access-to-water-ah247e00.pdf)
43 [content/uploads/2012/08/2006-Gomez-access-to-water-ah247e00.pdf](http://www.celep.info/wp-content/uploads/2012/08/2006-Gomez-access-to-water-ah247e00.pdf)
- 44 Gonzalez-Martin, C., Teigell-Perez, N., Valladares, B., & Griffin, D. W. (2014). *The Global*
45 *Dispersion of Pathogenic Microorganisms by Dust Storms and Its Relevance to Agriculture.*
46 *Advances in Agronomy* (Vol. 127). <https://doi.org/10.1016/B978-0-12-800131-8.00001-7>
- 47 Goudarzi, G., Daryanoosh, S. M., Godini, H., Hopke, P. K., Sicard, P., De Marco, A., ... Omidi

- 1 Khaniabadi, Y. (2017). Health risk assessment of exposure to the Middle-Eastern Dust storms in
2 the Iranian megacity of Kermanshah. *Public Health*, 148, 109–116.
3 <https://doi.org/10.1016/J.PUHE.2017.03.009>
- 4 Goudie, A. (2013). *The human impact on the natural environment : past, present and future* (7th ed.).
5 Wiley-Blackwell. Retrieved from
6 <https://books.google.de/books?hl=en&lr=&id=TxfRZr4nYrMC&oi=fnd&pg=PT12&dq=dust+s>
7 [kin+irritations++goudie&ots=ciZ1mfBKyN&sig=QBxgSSkl2NbrprxYZPRlccgVrU#v=onepag](https://books.google.de/books?hl=en&lr=&id=TxfRZr4nYrMC&oi=fnd&pg=PT12&dq=dust+s)
8 [e&q&f=false](https://books.google.de/books?hl=en&lr=&id=TxfRZr4nYrMC&oi=fnd&pg=PT12&dq=dust+s)
- 9 Goudie, A., & Middleton, N. (2006). *Desert dust in the global system*. Berlin ; Heidelberg; New York:
10 Springer. Retrieved from [https://www.worldcat.org/title/desert-dust-in-the-global-system-with-](https://www.worldcat.org/title/desert-dust-in-the-global-system-with-41-tables/oclc/180900043&referer=brief_results)
11 [41-tables/oclc/180900043&referer=brief_results](https://www.worldcat.org/title/desert-dust-in-the-global-system-with-41-tables/oclc/180900043&referer=brief_results)
- 12 Goudie, A. S. (2014). Desert dust and human health disorders. *Environment International*, 63, 101–
13 113. <https://doi.org/10.1016/J.ENVINT.2013.10.011>
- 14 Goudie, A. S., & Middleton, N. J. (2001). Saharan dust storms : nature and consequences. *Earth-*
15 *Science Reviews*, 56(1–4), 179–204. Retrieved from
16 <https://www.sciencedirect.com/science/article/pii/S0012825201000678>
- 17 Grace, J. B., M. D. Smith, S. L. Grace, S. L. Collins, and T. J. S. (2001). Interactions between fire
18 and invasive plants in temperate grasslands of North America. In K. E. M. Galley and T. P
19 Wilson (Ed.), *Proceedings of the invasive species workshop: the role of fire in the control and*
20 *spread of invasive species* (pp. 40–65). Tallahassee, Florida: Tall Timbers Research Station.
21 Retrieved from <https://www.researchgate.net/publication/239929069>
- 22 Grainger, A. (2009). The role of science in implementing international environmental agreements:
23 The case of desertification. *Land Degradation & Development*, 20(4), 410–430.
24 <https://doi.org/10.1002/ldr.898>
- 25 Gray, C. L. (2009). Environment, Land, and Rural Out-migration in the Southern Ecuadorian Andes.
26 *World Development*, 37(2), 457–468. <https://doi.org/10.1016/J.WORLDDEV.2008.05.004>
- 27 Gray, C. L. (2010). Gender, Natural Capital, and Migration in the Southern Ecuadorian Andes.
28 *Environment and Planning A*, 42(3), 678–696. <https://doi.org/10.1068/a42170>
- 29 Gray, C. L. (2011). Soil quality and human migration in Kenya and Uganda. *Global Environmental*
30 *Change*, 21(2), 421–430. <https://doi.org/10.1016/J.GLOENVCHA.2011.02.004>
- 31 Gray, C. L., & Bilsborrow, R. E. (2014). Consequences of out-migration for land use in rural Ecuador.
32 *Land Use Policy*, 36, 182–191. <https://doi.org/10.1016/J.LANDUSEPOL.2013.07.006>
- 33 Gray, C. L., & Mueller, V. (2012). Natural disasters and population mobility in Bangladesh.
34 *Proceedings of the National Academy of Sciences of the United States of America*, 109(16),
35 6000–5. <https://doi.org/10.1073/pnas.1115944109>
- 36 Gray, C., & Mueller, V. (2012). Drought and Population Mobility in Rural Ethiopia. *World*
37 *Development*, 40(1), 134–145. <https://doi.org/10.1016/J.WORLDDEV.2011.05.023>
- 38 Greene, R., Timms, W., Rengasamy, P., Arshad, M., & Cresswell, R. (2016). Soil and Aquifer
39 Salinization: Toward an Integrated Approach for Salinity Management of Groundwater. In
40 *Integrated Groundwater Management* (pp. 377–412). Cham: Springer International Publishing.
41 https://doi.org/10.1007/978-3-319-23576-9_15
- 42 Greve, P., Orlowsky, B., Mueller, B., Sheffield, J., Reichstein, M., & Seneviratne, S. I. (2014). Global
43 assessment of trends in wetting and drying over land. *Nature Geoscience*, 7(10), 716–721.
44 <https://doi.org/10.1038/ngeo2247>
- 45 Greve, P., Roderick, M. L., & Seneviratne, S. I. (2017). Simulated changes in aridity from the last
46 glacial maximum to 4xCO₂. *Environmental Research Letters*, 12(11), 114021.
47 <https://doi.org/10.1088/1748-9326/aa89a3>

- 1 Greve, P., & Seneviratne, S. I. (1999). Assessment of future changes in water availability and aridity.
2 *Geophysical Research Letters*, 26, 5493–5499. <https://doi.org/10.1002/2015GL064127>. Received
- 3 Griffin, D. W. (2007). Atmospheric movement of microorganisms in clouds of desert dust and
4 implications for human health. *Clinical Microbiology Reviews*, 20(3), 459–477.
5 <https://doi.org/10.1128/CMR.00039-06>
- 6 Griffin, D. W., Westphal, D. L., & Gray, M. A. (2006). Airborne microorganisms in the African
7 desert dust corridor over the mid-Atlantic ridge, Ocean Drilling Program, Leg 209.
8 *Aerobiologia*, 22(3), 211–226. <https://doi.org/10.1007/s10453-006-9033-z>
- 9 Groneman, A. (1968). The soils of the wind erosion control camp area Karapinar, Turkey. Retrieved
10 from <https://library.wur.nl/WebQuery/wurpubs/fulltext/303656>
- 11 Guan, C., Li, X., Zhang, P., & Li, C. (2018). Effect of global warming on soil respiration and
12 cumulative carbon release in biocrust-dominated areas in the Tengger Desert, northern China.
13 *Journal of Soils and Sediments*, 1–10. <https://doi.org/10.1007/s11368-018-2171-4>
- 14 Guerra, C. A., Maes, J., Geijzendorffer, I., & Metzger, M. J. (2016). An assessment of soil erosion
15 prevention by vegetation in Mediterranean Europe: Current trends of ecosystem service
16 provision, 60, 213–222. <https://doi.org/10.1016/j.ecolind.2015.06.043>
- 17 Guerrero-Baena, M., Gómez-Limón, J., Guerrero-Baena, M. D., & Gómez-Limón, J. A. (2019).
18 Insuring Water Supply in Irrigated Agriculture: A Proposal for Hydrological Drought Index-
19 Based Insurance in Spain. *Water*, 11(4), 686. <https://doi.org/10.3390/w11040686>
- 20 Gunin, P., & Pankova, E. (2004). *Contemporary Processes of Degradation and Desertification of*
21 *Ecosystems of the East Asian Sector of Steppes and Forest-Steppes" In Modern Global Changes*
22 *of the Natural Environment*. Moscow: Scientific World.
- 23 Guodaar, L., Asante, F., Eshun, G., Abass, K., Afriyie, K., Appiah, D. O., ... Kpenekuu, F. (2019).
24 How do climate change adaptation strategies result in unintended maladaptive outcomes?
25 Perspectives of tomato farmers. *International Journal of Vegetable Science*, 1–17.
26 <https://doi.org/10.1080/19315260.2019.1573393>
- 27 Gupta, A. K., Dey, A. R., Shinde, C., Mahanta, H., Patel, C., Patel, R., ... Tole, P. (2016). Theory of
28 open inclusive innovation for reciprocal, responsive and respectful outcomes: coping creatively
29 with climatic and institutional risks. *Journal of Open Innovation: Technology, Market, and*
30 *Complexity*, 2(1), 16. <https://doi.org/10.1186/s40852-016-0038-8>
- 31 Gupta, R., Kienzler, K., Martius, C., Mirzabaev, A., Oweis, T., De Pauw, E., ... Thomas, R. (2009).
32 Research prospectus: a vision for sustainable land management research in Central Asia.
33 *ICARDA Central Asia and Caucasus Program. Sustainable Agriculture in Central Asia and the*
34 *Caucasus Series, 1*, 84.
- 35 Gurgung, J. D., Mwanundundu, S., Lubbock, A., Hartl, M., and Firmian, I., & IFAD. (2006). *Gender and*
36 *Desertification: Expanding Roles for Women to Restore Dryland Areas*. Rome: International
37 Fund for Agricultural Development (IFAD). Rome. Retrieved from
38 <https://www.ifad.org/documents/10180/02b5cc0c-5031-43d4-8ebe-7c23568bec80>
- 39 Gutiérrez-Elorza, M. (2006). Erosión e influencia del cambio climático. *Revista C & G*, 20(3–4), 45–
40 59.
- 41 Guzman, C. D., Tilahun, S. A., Dagnaw, D. C., Zegeye, A. D., Yitafaru, B., Kay, R. W., & Steenhuis,
42 T. S. (2018). Developing Soil Conservation Strategies with Technical and Community
43 Knowledge in a Degrading Sub-Humid Mountainous Landscape. *Land Degradation &*
44 *Development*, 29(3), 749–764. <https://doi.org/10.1002/ldr.2733>
- 45 Haan, C. De, Dubern, E., Garancher, B., & Quintero, C. (2016). *Pastoralism Development in the*
46 *Sahel: A Road to Stability?* World Bank Group. Retrieved from
47 <https://elibrary.worldbank.org/doi/abs/10.1596/24228>
- 48 Haase, P., Tonkin, J. D., Stoll, S., Burkhard, B., Frenzel, M., Geijzendorffer, I. R., ... Schmeller, D. S.

- 1 (2018). The next generation of site-based long-term ecological monitoring: Linking essential
2 biodiversity variables and ecosystem integrity. *Science of The Total Environment*, 613–614,
3 1376–1384. <https://doi.org/10.1016/J.SCITOTENV.2017.08.111>
- 4 Haggblade, S., Hazell, P., & Reardon, T. (2010). The Rural Non-farm Economy: Prospects for
5 Growth and Poverty Reduction. *World Development*, 38(10), 1429–1441.
6 <https://doi.org/10.1016/J.WORLDDEV.2009.06.008>
- 7 Hallegatte, S. (2012). *A Cost Effective Solution to Reduce Disaster Losses in Developing Countries:*
8 *Hydro-Meteorological Services, Early Warning, and Evacuation*. The World Bank.
9 <https://doi.org/10.1596/1813-9450-6058>
- 10 Hallegatte, S., Bangalore, M., Bonzanigo, L., Fay, M., Kane, T., Narloch, U., ... Vogt-Schilb, A.
11 (2015). *Shock Waves: Managing the Impacts of Climate Change on Poverty*. The World Bank.
12 <https://doi.org/10.1596/978-1-4648-0673-5>
- 13 Hallegatte, S., & Rozenberg, J. (2017). Climate change through a poverty lens. *Nature Climate*
14 *Change*, 7(4), 250–256. <https://doi.org/10.1038/nclimate3253>
- 15 Halliday, T. (2006). Migration, Risk, and Liquidity Constraints in El Salvador. *Economic*
16 *Development and Cultural Change*, 54(4), 893–925. <https://doi.org/10.1086/503584>
- 17 Hamidov, A., Helming, K., & Balla, D. (2016). Impact of agricultural land use in Central Asia: a
18 review. *Agronomy for Sustainable Development*, 36(1), 6. [https://doi.org/10.1007/s13593-015-](https://doi.org/10.1007/s13593-015-0337-7)
19 [0337-7](https://doi.org/10.1007/s13593-015-0337-7)
- 20 Han, G., Hao, X., Zhao, M., Wang, M., Ellert, B. H., Willms, W., & Wang, M. (2008). Effect of
21 grazing intensity on carbon and nitrogen in soil and vegetation in a meadow steppe in Inner
22 Mongolia. *Agriculture, Ecosystems & Environment*, 125(1), 21–32.
23 <https://doi.org/https://doi.org/10.1016/j.agee.2007.11.009>
- 24 Han, Q., Luo, G., Li, C., Shakir, A., Wu, M., & Saidov, A. (2016). Simulated grazing effects on
25 carbon emission in Central Asia. *Agricultural and Forest Meteorology*, 216, 203–214.
26 <https://doi.org/10.1016/J.AGRFORMET.2015.10.007>
- 27 Hanjra, M. A., Blackwell, J., Carr, G., Zhang, F., & Jackson, T. M. (2012). Wastewater irrigation and
28 environmental health: Implications for water governance and public policy. *International*
29 *Journal of Hygiene and Environmental Health*, 215(3), 255–269.
30 <https://doi.org/10.1016/J.IJHEH.2011.10.003>
- 31 Hao, Y., Wang, Y., Mei, X., Huang, X., Cui, X., Zhou, X., & Niu, H. (2008). CO₂, H₂O and energy
32 exchange of an Inner Mongolia steppe ecosystem during a dry and wet year. *Acta Oecologica*,
33 33(2), 133–143. <https://doi.org/10.1016/j.actao.2007.07.002>
- 34 Harou, J. J., Medellín-Azuara, J., Zhu, T., Tanaka, S. K., Lund, J. R., Stine, S., ... Jenkins, M. W.
35 (2010). Economic consequences of optimized water management for a prolonged, severe
36 drought in California. *Water Resources Research*, 46(5).
37 <https://doi.org/10.1029/2008WR007681>
- 38 Harpold, A. A., & Brooks, P. D. (2018). Humidity determines snowpack ablation under a warming
39 climate. *Proceedings of the National Academy of Sciences of the United States of America*,
40 115(6), 1215–1220. <https://doi.org/10.1073/pnas.1716789115>
- 41 Harris, D., & Orr, A. (2014). Is rainfed agriculture really a pathway from poverty? *Agricultural*
42 *Systems*, 123, 84–96. <https://doi.org/10.1016/J.AGSY.2013.09.005>
- 43 Hartmann, B. (2010). Rethinking climate refugees and climate conflict: Rhetoric, reality and the
44 politics of policy discourse. *Journal of International Development*, 22(2), 233–246.
45 <https://doi.org/10.1002/jid.1676>
- 46 Hasanean, H., Almazroui, M., Hasanean, H., & Almazroui, M. (2015). Rainfall: Features and
47 Variations over Saudi Arabia, A Review. *Climate*, 3(3), 578–626.
48 <https://doi.org/10.3390/cli3030578>

- 1 Havstad, K. M., L. F. Huenneke, and W. H. S. (2006). *Structure and Function of a Chihuahuan*
2 *Desert Ecosystem: The Jornada Basin ...* - Google Books. Oxford University Press. Retrieved
3 from [https://books.google.co.ke/books?hl=en&lr=&id=-](https://books.google.co.ke/books?hl=en&lr=&id=-T8PpZ5Xn4wC&oi=fnd&pg=PR11&dq=Havstad,+K.+M.,+L.+F.+Huenneke,+and+W.+H.+Schlesinger+%5Beds.%5D.+2006.+Structure+and+function+of+a+Chihuahuan+Desert+ecosystem:+the+Jornada+Basin+long-term+ecological+research+s)
4 [T8PpZ5Xn4wC&oi=fnd&pg=PR11&dq=Havstad,+K.+M.,+L.+F.+Huenneke,+and+W.+H.+Sch](https://books.google.co.ke/books?hl=en&lr=&id=-T8PpZ5Xn4wC&oi=fnd&pg=PR11&dq=Havstad,+K.+M.,+L.+F.+Huenneke,+and+W.+H.+Schlesinger+%5Beds.%5D.+2006.+Structure+and+function+of+a+Chihuahuan+Desert+ecosystem:+the+Jornada+Basin+long-term+ecological+research+s)
5 [lesinger+%5Beds.%5D.+2006.+Structure+and+function+of+a+Chihuahuan+Desert+ecosystem:](https://books.google.co.ke/books?hl=en&lr=&id=-T8PpZ5Xn4wC&oi=fnd&pg=PR11&dq=Havstad,+K.+M.,+L.+F.+Huenneke,+and+W.+H.+Schlesinger+%5Beds.%5D.+2006.+Structure+and+function+of+a+Chihuahuan+Desert+ecosystem:+the+Jornada+Basin+long-term+ecological+research+s)
6 [+the+Jornada+Basin+long-term+ecological+research+s](https://books.google.co.ke/books?hl=en&lr=&id=-T8PpZ5Xn4wC&oi=fnd&pg=PR11&dq=Havstad,+K.+M.,+L.+F.+Huenneke,+and+W.+H.+Schlesinger+%5Beds.%5D.+2006.+Structure+and+function+of+a+Chihuahuan+Desert+ecosystem:+the+Jornada+Basin+long-term+ecological+research+s)
- 7 He, J., & Lang, R. (2015). Limits of State-Led Programs of Payment for Ecosystem Services: Field
8 Evidence from the Sloping Land Conversion Program in Southwest China. *Human Ecology*,
9 *43*(5), 749–758. <https://doi.org/10.1007/s10745-015-9782-9>
- 10 He, J., & Sikor, T. (2015). Notions of justice in payments for ecosystem services: Insights from
11 China's Sloping Land Conversion Program in Yunnan Province. *Land Use Policy*, *43*, 207–216.
12 <https://doi.org/10.1016/J.LANDUSEPOL.2014.11.011>
- 13 Headey, D., Taffesse, A. S., & You, L. (2014). Diversification and Development in Pastoralist
14 Ethiopia. *World Development*, *56*, 200–213.
15 <https://doi.org/10.1016/J.WORLDDEV.2013.10.015>
- 16 Hein, L., & De Ridder, N. (2006). Desertification in the Sahel: a reinterpretation. *Global Change*
17 *Biology*, *12*(5), 751–758. <https://doi.org/10.1111/j.1365-2486.2006.01135.x>
- 18 Helldén, U., & Tottrup, C. (2008). Regional desertification: A global synthesis. *Global and Planetary*
19 *Change*, *64*(3–4), 169–176. <https://doi.org/10.1016/j.gloplacha.2008.10.006>
- 20 Hellmann, J. J., Byers, J. E., Bierwagen, B. G., & Dukes, J. S. (2008). Five Potential Consequences of
21 Climate Change for Invasive Species. *Conservation Biology*, *22*(3), 534–543.
22 <https://doi.org/10.1111/j.1523-1739.2008.00951.x>
- 23 Henderson, J. V., Storeygard, A., & Deichmann, U. (2017). Has climate change driven urbanization in
24 Africa? *Journal of Development Economics*, *124*, 60–82.
25 <https://doi.org/10.1016/J.JDEVCO.2016.09.001>
- 26 Henry, S., Schoumaker, B., & Beauchemin, C. (2004). The Impact of Rainfall on the First Out-
27 Migration: A Multi-level Event-History Analysis in Burkina Faso. *Population and Environment*,
28 *25*(5), 423–460. <https://doi.org/10.1023/B:POEN.0000036928.17696.e8>
- 29 Hernandez, R. R., & Sandquist, D. R. (2011). Disturbance of biological soil crust increases emergence
30 of exotic vascular plants in California sage scrub. *Plant Ecology*, *212*(10), 1709–1721.
31 <https://doi.org/10.1007/s11258-011-9943-x>
- 32 Herrera, C., Gamboa, C., Custodio, E., Jordan, T., Godfrey, L., Jódar, J., ... Sáez, A. (2018). Science
33 of the Total Environment Groundwater origin and recharge in the hyperarid Cordillera de la
34 Costa , Atacama Desert , northern Chile. *Science of the Total Environment*, *624*, 114–132.
35 <https://doi.org/10.1016/j.scitotenv.2017.12.134>
- 36 Herrero, S. T. (2006). Desertification and Environmental Security: The Case of Conflicts between
37 Farmers and Herders in the Arid Environments of the Sahel. In *Desertification in the*
38 *Mediterranean Region. A Security Issue* (pp. 109–132). Dordrecht: Kluwer Academic
39 Publishers. https://doi.org/10.1007/1-4020-3760-0_04
- 40 Herrick, J. E., Karl, J. W., McCord, S. E., Buenemann, M., Riginos, C., Courtright, E., ...
41 Bestelmeyer, B. (2017). Two New Mobile Apps for Rangeland Inventory and Monitoring by
42 Landowners and Land Managers. *Rangelands*, *39*(2), 46–55.
43 <https://doi.org/10.1016/j.rala.2016.12.003>
- 44 Herrmann, S. M., & Hutchinson, C. F. (2005). The changing contexts of the desertification debate.
45 *Journal of Arid Environments*, *63*(3), 538–555.
46 <https://doi.org/10.1016/J.JARIDENV.2005.03.003>
- 47 Herrmann, S. M., & Sop, T. K. (2016). The Map Is not the Territory: How Satellite Remote Sensing
48 and Ground Evidence Have Re-shaped the Image of Sahelian Desertification (pp. 117–145).

- 1 Springer, Berlin, Heidelberg. https://doi.org/10.1007/978-3-642-16014-1_5
- 2 Herrmann, S. M., & Tappan, G. G. (2013). Vegetation impoverishment despite greening: A case study
3 from central Senegal. *Journal of Arid Environments*, *90*, 55–66.
4 <https://doi.org/10.1016/J.JARIDENV.2012.10.020>
- 5 Hertel, T. W., Burke, M. B., & Lobell, D. B. (2010). The poverty implications of climate-induced
6 crop yield changes by 2030. *Global Environmental Change*, *20*(4), 577–585.
7 <https://doi.org/10.1016/J.GLOENVCHA.2010.07.001>
- 8 Hertel, T. W., & Lobell, D. B. (2014). Agricultural adaptation to climate change in rich and poor
9 countries: Current modeling practice and potential for empirical contributions. *Energy*
10 *Economics*, *46*, 562–575. <https://doi.org/10.1016/J.ENECO.2014.04.014>
- 11 Hetem, R. S., Maloney, S. K., Fuller, A., & Mitchell, D. (2016). Heterothermy in large mammals:
12 inevitable or implemented? *Biological Reviews*, *91*(1), 187–205.
13 <https://doi.org/10.1111/brv.12166>
- 14 Hiernaux, P., Ayantunde, A., Kalilou, A., Mougin, E., Gérard, B., Baup, F., ... Djaby, B. (2009).
15 Trends in productivity of crops, fallow and rangelands in Southwest Niger: Impact of land use,
16 management and variable rainfall. *Journal of Hydrology*, *375*(1–2), 65–77.
17 <https://doi.org/10.1016/J.JHYDROL.2009.01.032>
- 18 Hiernaux, P., & Houérou, H. N. Le. (2006). *Sécheresse. Science et changements planétaires /*
19 *Sécheresse* (Vol. 17). John Libbey Eurotext Ltd. Retrieved from
20 http://www.jle.com/fr/revues/sec/e-docs/les_parours_du_sahel_270087/article.phtml
- 21 Higginbottom, T., & Symeonakis, E. (2014). Assessing Land Degradation and Desertification Using
22 Vegetation Index Data: Current Frameworks and Future Directions. *Remote Sensing*, *6*(10),
23 9552–9575. <https://doi.org/10.3390/rs6109552>
- 24 Higgins, D., Balint, T., Liversage, H., & Winters, P. (2018). Investigating the impacts of increased
25 rural land tenure security: A systematic review of the evidence. *Journal of Rural Studies*, *61*,
26 34–62. <https://doi.org/10.1016/J.JRURSTUD.2018.05.001>
- 27 Hijioka, Y., E. Lin, J.J. Pereira, R.T. Corlett, X. Cui, G.E. Insarov, R.D. Lasco, E. Lindgren, and A.
28 S. (2014). Asia. In K. L. E. Barros, V.R., C.B. Field, D.J. Dokken, M.D. Mastrandrea, K.J.
29 Mach, T.E. Bilir, M. Chatterjee & and L. L. W. Y.O. Estrada, R.C. Genova, B. Girma, E.S.
30 Kissel, A.N. Levy, S. MacCracken, P.R. Mastrandrea (Eds.), *Climate Change 2014: Impacts,*
31 *Adaptation, and Vulnerability. Part B: Regional Aspects. Contribution of Working Group II to*
32 *the Fifth Assessment Report of the Intergovernmental Panel on Climate Change* (pp. 1327–1).
33 Cambridge, United Kingdom and New York, NY, USA: Cambridge University Press. Retrieved
34 from https://www.ipcc.ch/site/assets/uploads/2018/02/WGIIAR5-Chap24_FINAL.pdf
- 35 Hirche, A., Salamani, M., Abdellaoui, A., Benhouhou, S., & Valderrama, J. M. (2011). Landscape
36 changes of desertification in arid areas: the case of south-west Algeria. *Environmental*
37 *Monitoring and Assessment*, *179*(1–4), 403–420. <https://doi.org/10.1007/s10661-010-1744-5>
- 38 Hirche, A., Salamani, M., Tarhouni, M., Nedjraoui, D., Gashut, E. H., Abdelmadjid, B., ... Amrani, S.
39 (2018). The Maghreb (North Africa) rangelands evolution for forty years: Regreening or
40 degradation? In *Desertification: Past, Current and Future Trends* (pp. 73–106). Nova.
41 Retrieved from <https://www.researchgate.net/publication/328007558>
- 42 Hochman, Z., Gobbett, D. L., & Horan, H. (2017). Climate trends account for stalled wheat yields in
43 Australia since 1990. *Global Change Biology*, *23*(5), 2071–2081.
44 <https://doi.org/10.1111/gcb.13604>
- 45 Hoegh-Guldberg, O., Jacob, D., Taylor, M., Bindi, M., Abdul Halim, S., Achlatis Australia, M., ...
46 Waterfield, T. (2018). IPCCchapter3.
- 47 Hoegh-Guldberg, O., Jacob, D., Taylor, M., Bindi, M., Brown, S., Camilloni, I., ... IPCC. (2018).
48 Chapter 3: Impacts of 1.5°C global warming on natural and human systems. In *GLOBAL*

- 1 *WARMING OF 1.5 °C*. IPCC. Retrieved from http://report.ipcc.ch/sr15/pdf/sr15_chapter3.pdf
- 2 Hoerling, M., Hurrell, J., Eischeid, J., Phillips, A., Hoerling, M., Hurrell, J., ... Phillips, A. (2006).
3 Detection and Attribution of Twentieth-Century Northern and Southern African. *Climate*
4 *Research*, 19(16), 3989–4008. <https://doi.org/10.1175/JCLI3842.1>
- 5 Hoffmann, U., Yair, A., Hikel, H., & Kuhn, N. J. (2012). Soil organic carbon in the rocky desert of
6 northern Negev (Israel). *Journal of Soils and Sediments*, 12(6), 811–825.
7 <https://doi.org/10.1007/s11368-012-0499-8>
- 8 Holden, S. R., Gutierrez, A., & Treseder, K. K. (2013). Changes in Soil Fungal Communities,
9 Extracellular Enzyme Activities, and Litter Decomposition Across a Fire Chronosequence in
10 Alaskan Boreal Forests. *Ecosystems*, 16(1), 34–46. <https://doi.org/10.1007/s10021-012-9594-3>
- 11 Holden, S. T., & Ghebru, H. (2016). Land tenure reforms, tenure security and food security in poor
12 agrarian economies: Causal linkages and research gaps. *Global Food Security*, 10, 21–28.
13 <https://doi.org/10.1016/J.GFS.2016.07.002>
- 14 Hong, L., Zhou, N., Fridley, D., & Raczkowski, C. (2013). Assessment of China's renewable energy
15 contribution during the 12th Five Year Plan. *Energy Policy*, 62, 1533–1543.
16 <https://doi.org/10.1016/J.ENPOL.2013.07.110>
- 17 Hopke, P. K., Saeed, S., Nazari, H., Hadei, M., Yarahmadi, M., Kermani, M., ... Shahsavani, A.
18 (2018). Spatial and Temporal Trends of Short-Term Health Impacts of PM 2.5 in Iranian Cities;
19 a Modelling Approach (2013-2016). *Aerosol and Air Quality Research*, 18, 497–504.
20 <https://doi.org/10.4209/aaqr.2017.09.0325>
- 21 Hopkins, A., & Del Prado, A. (2007). Implications of climate change for grassland in Europe:
22 impacts, adaptations and mitigation options: a review. *Grass and Forage Science*, 62(2), 118–
23 126. <https://doi.org/10.1111/j.1365-2494.2007.00575.x>
- 24 Horion, S., Prishchepov, A. V., Verbesselt, J., de Beurs, K., Tagesson, T., & Fensholt, R. (2016).
25 Revealing turning points in ecosystem functioning over the Northern Eurasian agricultural
26 frontier. *Global Change Biology*, 22(8), 2801–2817. <https://doi.org/10.1111/gcb.13267>
- 27 Hornbeck, R. (2012). The Enduring Impact of the American Dust Bowl: Short- and Long-Run
28 Adjustments to Environmental Catastrophe. *American Economic Review*, 102(4), 1477–1507.
29 <https://doi.org/10.1257/aer.102.4.1477>
- 30 Houyou, Z., Bienders, C. L., Benhorma, H. A., Dellal, A., & Boutemdjet, A. (2016). Evidence of
31 Strong Land Degradation by Wind Erosion as a Result of Rainfed Cropping in the Algerian
32 Steppe: A Case Study at Laghouat. *Land Degradation & Development*, 27(8), 1788–1796.
33 <https://doi.org/10.1002/ldr.2295>
- 34 Howe, C., Suich, H., van Gardingen, P., Rahman, A., & Mace, G. M. (2013). Elucidating the
35 pathways between climate change, ecosystem services and poverty alleviation. *Current Opinion*
36 *in Environmental Sustainability*, 5(1), 102–107. <https://doi.org/10.1016/J.COSUST.2013.02.004>
- 37 Hoyle, F. C., D'Antuono, M. D., Overheu, T., & Murphy, D. V. (2013). Capacity for increasing soil
38 organic carbon stocks in dryland agricultural systems. *Soil Research*, 51, 657–667.
- 39 Hsiang, S., Burke, M., & E Miguel -. (2013). Quantifying the influence of climate on human conflict.
40 *Science*, 341(6151). Retrieved from
41 <http://science.sciencemag.org/content/341/6151/1235367.short>
- 42 Hu, Y., & Fu, Q. (2007). Observed poleward expansion of the Hadley circulation since 1979.
43 *Atmospheric Chemistry and Physics*, 7(19), 5229–5236. [https://doi.org/10.5194/acp-7-5229-](https://doi.org/10.5194/acp-7-5229-2007)
44 [2007](https://doi.org/10.5194/acp-7-5229-2007)
- 45 Huang, D., Wang, K., & Wu, W. L. (2007). Dynamics of soil physical and chemical properties and
46 vegetation succession characteristics during grassland desertification under sheep grazing in an
47 agro-pastoral transition zone in Northern China. *Journal of Arid Environments*, 70(1), 120–136.
48 <https://doi.org/10.1016/j.jaridenv.2006.12.009>

- 1 Huang, J., Fu, Q., Su, J., Tang, Q., Minnis, P., Hu, Y., ... Zhao, Q. (2009). Taklimakan dust aerosol
2 radiative heating derived from CALIPSO observations using the Fu-Liou radiation model with
3 CERES constraints. *Atmospheric Chemistry and Physics*, 9(12), 4011–4021.
4 <https://doi.org/10.5194/acp-9-4011-2009>
- 5 Huang, J., Li, Y., Fu, C., Chen, F., Fu, Q., Dai, A., ... Wang, G. (2017). Dryland climate change:
6 Recent progress and challenges. *Reviews of Geophysics*, 55(3), 719–778.
7 <https://doi.org/10.1002/2016RG000550>
- 8 Huang, J., Wang, T., Wang, W., Li, Z., & Yan, H. (2014). Climate effects of dust aerosols over East
9 Asian arid and semiarid regions. *Journal of Geophysical Research: Atmospheres*, 119(19),
10 11,398–11,416. <https://doi.org/10.1002/2014JD021796>
- 11 Huang, J., Yu, H., Dai, A., Wei, Y., & Kang, L. (2017). Drylands face potential threat under 2 °C
12 global warming target. *Nature Climate Change*, 7(6), 417–422.
13 <https://doi.org/10.1038/nclimate3275>
- 14 Huang, J., Yu, H., Guan, X., Wang, G., & Guo, R. (2016). Accelerated dryland expansion under
15 climate change. *Nature Climate Change*, 6(2), 166–171. <https://doi.org/10.1038/nclimate2837>
- 16 Huang, J., Yu, H., Guan, X., Wang, G., Guo, R., & Guo, R. (2016). Accelerated dryland expansion under
17 climate change. *Nature Climate Change*, 6(2), 166–171. <https://doi.org/10.1038/nclimate2837>
- 18 Huber, S., & Fensholt, R. (2011). Analysis of teleconnections between AVHRR-based sea surface
19 temperature and vegetation productivity in the semi-arid Sahel. *Remote Sensing of Environment*,
20 115(12), 3276–3285. <https://doi.org/10.1016/J.RSE.2011.07.011>
- 21 Huber, S., Fensholt, R., & Rasmussen, K. (2011). Water availability as the driver of vegetation
22 dynamics in the African Sahel from 1982 to 2007. *Global and Planetary Change*, 76(3–4), 186–
23 195. <https://doi.org/10.1016/J.GLOPLACHA.2011.01.006>
- 24 Hudak, A. T., Wessman, C. A., & Seastedt, T. R. (2003). Woody overstorey effects on soil carbon and
25 nitrogen pools in a SouthAfrican savanna. *Austral Ecology*, 28, 173–181.
- 26 Hughes, R. F., Archer, S. R., Asner, G. P., Wessman, C. A., McMurtry, C., Nelson, J., & Ansley, R. J.
27 (2006). Changes in aboveground primary production and carbon and nitrogen pools
28 accompanying woody plant encroachment in a temperate savanna. *Global Change Biology*,
29 12(9), 1733–1747. <https://doi.org/10.1111/j.1365-2486.2006.01210.x>
- 30 Hugo, G. (2008). *Migration, development and environment*. International Organization for Migration.
31 Retrieved from http://www.ciesin.columbia.edu/repository/pern/papers/hugo_statement.pdf
- 32 Hull, A. C., & Stewart, G. (1949). *Agronomy journal*. *Journal of the American Society of Agronomy*.
33 American Society of Agronomy. Retrieved from <http://agris.fao.org/agris-search/search.do?recordID=US201301502209>
- 35 Hunter, L. M., Luna, J. K., & Norton, R. M. (2015). Environmental Dimensions of Migration. *Annual*
36 *Review of Sociology*, 41(1), 377–397. <https://doi.org/10.1146/annurev-soc-073014-112223>
- 37 Hurlbert, M. A. (2018). *Adaptive Governance of Disaster: Drought and Flood in Rural Areas*. Cham,
38 Switzerland: Springer.
- 39 Hurlbert, M., & Mussetta, P. (2016). Creating resilient water governance for irrigated producers in
40 Mendoza, Argentina. *Environmental Science and Policy*, 58, 83–94.
41 <https://doi.org/10.1016/j.envsci.2016.01.004>
- 42 Hussein, I. A. El. (2011). Desertification Process in Egypt (pp. 863–874). https://doi.org/10.1007/978-3-642-17776-7_50
- 44 Ibragimov, N., Evett, S. R., Esanbekov, Y., Kamilov, B. S., Mirzaev, L., & Lamers, J. P. A. (2007).
45 Water use efficiency of irrigated cotton in Uzbekistan under drip and furrow irrigation.
46 *Agricultural Water Management*, 90(1), 112–120.
47 <https://doi.org/https://doi.org/10.1016/j.agwat.2007.01.016>

- 1 Ibrahim, H. I. . (2016). Tolerance to two Pomegranate cultivars (*Punica granatum* L.) to salinity stress
2 under hydroponic conditions. *J. Basic Appl. Sci. Res*, 6(4), 38–46. Retrieved from
3 [https://www.textroad.com/pdf/JBASR/J. Basic. Appl. Sci. Res., 6\(4\)38-46, 2016.pdf](https://www.textroad.com/pdf/JBASR/J. Basic. Appl. Sci. Res., 6(4)38-46, 2016.pdf)
- 4 Ibrahim, Y., Balzter, H., Kaduk, J., & Tucker, C. (2015). Land Degradation Assessment Using
5 Residual Trend Analysis of GIMMS NDVI3g, Soil Moisture and Rainfall in Sub-Saharan West
6 Africa from 1982 to 2012. *Remote Sensing*, 7(5), 5471–5494.
7 <https://doi.org/10.3390/rs70505471>
- 8 IEA. (2013). *World Energy Outlook 2013*. Retrieved from <http://www.iea.org/>
- 9 Ifejika Speranza, C., Kiteme, B., Ambenje, P., Wiesmann, U., & Makali, S. (2010). Indigenous
10 knowledge related to climate variability and change: insights from droughts in semi-arid areas of
11 former Makueni District, Kenya. *Climatic Change*, 100(2), 295–315.
12 <https://doi.org/10.1007/s10584-009-9713-0>
- 13 Iiyama, M., Dobie, P., Njenga, M., Ndegwa, G., & Jamnadass, R. (2014). The potential of
14 agroforestry in the provision of sustainable woodfuel in sub-Saharan Africa. *Current Opinion in*
15 *Environmental Sustainability*, 6, 138–147. <https://doi.org/10.1016/J.COSUST.2013.12.003>
- 16 Iknayan, K. J., & Beissinger, S. R. (2018). Collapse of a desert bird community over the past century
17 driven by climate change. *Proceedings of the National Academy of Sciences of the United States*
18 *of America*, 115(34), 8597–8602. <https://doi.org/10.1073/pnas.1805123115>
- 19 Imeson, A. (2012). *Desertification, land degradation and sustainability*. Wiley-Blackwell. Retrieved
20 from
21 [https://books.google.com/books?hl=en&lr=&id=ZcquTQR9g1sC&oi=fnd&pg=PR11&dq=Imeson,+A.+\(2012\).+Desertification+Indicators:+From+Concept+to+Practice.+In+Desertification,+Land+Degradation+and+Sustainability,+A.+Imeson+\(Ed.\)&ots=nBo7bmaNvB&sig=Ad0Wk9VaOn5ApEX](https://books.google.com/books?hl=en&lr=&id=ZcquTQR9g1sC&oi=fnd&pg=PR11&dq=Imeson,+A.+(2012).+Desertification+Indicators:+From+Concept+to+Practice.+In+Desertification,+Land+Degradation+and+Sustainability,+A.+Imeson+(Ed.)&ots=nBo7bmaNvB&sig=Ad0Wk9VaOn5ApEX)
- 25 Inbar, M. (2007). Importance of Drought Information in Monitoring and Assessing Land Degradation.
26 In *Climate and Land Degradation* (pp. 253–266). Berlin, Heidelberg: Springer Berlin
27 Heidelberg. https://doi.org/10.1007/978-3-540-72438-4_13
- 28 Indoitu, R., Kozhoridze, G., Batyrbaeva, M., Vitkovskaya, I., Orlovsky, N., Blumberg, D., &
29 Orlovsky, L. (2015). Dust emission and environmental changes in the dried bottom of the Aral
30 Sea. *Aeolian Research*, 17, 101–115. <https://doi.org/10.1016/J.AEOLIA.2015.02.004>
- 31 Infante, F. (2017). The role of social capital and labour exchange in the soils of Mediterranean Chile.
32 *Rural Society*, 26(2), 107–124. <https://doi.org/10.1080/10371656.2017.1330837>
- 33 IPBES. (2018a). *Assessment Report on Land Degradation and Restoration | IPBES. Montanarella, L.,*
34 *Scholes, R., and Brainich, A. (eds.)*. Bonn. Retrieved from [https://www.ipbes.net/assessment-](https://www.ipbes.net/assessment-reports/ldr)
35 [reports/ldr](https://www.ipbes.net/assessment-reports/ldr)
- 36 IPBES. (2018b). *Summary for policymakers of the thematic assessment report on land degradation*
37 *and restoration of the Intergovernmental Science-Policy Platform on Biodiversity and*
38 *Ecosystem Services*. (J. F. R. Scholes, L. Montanarella, A. Brainich, N. Barger, B. ten Brink, M.
39 Cantele, B. Erasmus & M. S. and L. W. T. Gardner, T. G. Holland, F. Kohler, J. S. Kotiaho, G.
40 Von Maltitz, G. Nangendo, R. Pandit, J. Parrotta, M. D. Potts, S. Prince, Eds.). Bonn, Germany:
41 IPBES Secretariat.
- 42 IPCC. (2007). *Climate Change 2007: Impacts, Adaptation and Vulnerability. Contribution of Working*
43 *Group II to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change*.
44 (P. J. van der L. and C. E. M.L. Parry, O.F. Canziani, J.P. Palutikof & Hanson, Eds.). UK:
45 Cambridge University Press, Cambridge. Retrieved from [https://www.ipcc.ch/pdf/assessment-](https://www.ipcc.ch/pdf/assessment-report/ar4/wg2/ar4_wg2_full_report.pdf)
46 [report/ar4/wg2/ar4_wg2_full_report.pdf](https://www.ipcc.ch/pdf/assessment-report/ar4/wg2/ar4_wg2_full_report.pdf)
- 47 IPCC. (2012). *Managing the Risks of Extreme Events and Disasters to Advance Climate Change*
48 *Adaptation. A Special Report of Working Groups I and II of the Intergovernmental Panel on*
49 *Climate Change [Field, C.B., V. Barros, T.F. Stocker, D. Qin, D.J. Dokken, K.L. Ebi, M.D.*

- 1 IPCC. (2013). *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I to*
2 *the Fifth Assessment Report of the Intergovernmental Panel on Climate Change.* (T. F. Stocker,
3 D. Qin, G.-K. Plattner, M. Tignor, S. K. Allen, J. Boschung, ... P. M. Midgley, Eds.).
4 Cambridge University Press. Retrieved from [http://www.climatechange2013.org/report/full-](http://www.climatechange2013.org/report/full-report/)
5 [report/](http://www.climatechange2013.org/report/full-report/)
- 6 Irshad, M., Inoue, M., Ashraf, M., Faridullah, Delower, H. K. M., & Tsunekawa, A. (2007). Land
7 Desertification-An Emerging Threat to Environment and Food Security of Pakistan. *Journal of*
8 *Applied Sciences*, 7(8), 1199–1205. <https://doi.org/10.3923/jas.2007.1199.1205>
- 9 Islam, M. N., & Almazroui, M. (2012). Direct effects and feedback of desert dust on the climate of the
10 Arabian Peninsula during the wet season: a regional climate model study. *Climate Dynamics*,
11 39(9–10), 2239–2250. <https://doi.org/10.1007/s00382-012-1293-4>
- 12 Issanova, G., & Abuduwaili, J. (2017). Relationship Between Storms and Land Degradation (pp. 71–
13 86). https://doi.org/10.1007/978-981-10-3190-8_4
- 14 Ivanov, A. L., Kust, G. S., Kozlov, D. N., & and others. (2018). *National Report Global climate and*
15 *soil cover of Russia: assessment of risks and environmental and economic consequences of land*
16 *degradation. Adaptive systems and technologies of environmental management (agriculture and*
17 *forestry). In Russian: NATSIONAL'N. Moscow.*
- 18 Jackson, H., & Prince, S. D. (2016). Degradation of net primary production in a semiarid rangeland.
19 *Biogeosciences*, 13, 4721–4734. <https://doi.org/10.5194/bg-13-4721-2016>
- 20 Jackson, R. B., Banner, J. L., Jobbagy, E. G., Pockman, W. T., & Wall, D. H. (2002). Ecosystem
21 carbon loss with woody plant invasion of grasslands. *Nature*, 418(6898), 623–626. Retrieved
22 from <http://dx.doi.org/10.1038/nature00910>
- 23 Jan, S. U., Jamil, M., Alipour, H. A. D. I., Bhatti, M. F., & Gul, A. L. V. I. N. A., Jan, S. U., Jamil,
24 M., Alipour, H., Bhatti, M. F., & Gul, A. (2017). Analysis of salinity tolerance potential in
25 synthetic hexaploid wheat. *Pak. J. Bot*, 49(4), 1269–1278.
- 26 Jedwab, R., & Vollrath, D. (2015). Urbanization without growth in historical perspective.
27 *Explorations in Economic History*, 58, 1–21. <https://doi.org/10.1016/J.EEH.2015.09.002>
- 28 Jeon, E. M., Kim, H. J., Jung, K., Kim, J. H., Kim, M. Y., Kim, Y. P., & Ka, J. O. (2011). Impact of
29 Asian dust events on airborne bacterial community assessed by molecular analyses. *Atmospheric*
30 *Environment*, 45(25), 4313–4321. <https://doi.org/10.1016/j.atmosenv.2010.11.054>
- 31 Jerneck, A., & Olsson, L. (2008). Adaptation and the poor: development, resilience and transition.
32 *Climate Policy*, 8(2), 170–182. <https://doi.org/10.3763/cpol.2007.0434>
- 33 Ji, M., Huang, J., Xie, Y., & Liu, J. (2015). Comparison of dryland climate change in observations
34 and CMIP5 simulations. *Advances in Atmospheric Sciences*, 32(11), 1565–1574.
35 <https://doi.org/10.1007/s00376-015-4267-8>
- 36 Jiang, H. (2016). Taking Down the “Great Green Wall”: The Science and Policy Discourse of
37 Desertification and Its Control in China (pp. 513–536). Springer, Berlin, Heidelberg.
38 https://doi.org/10.1007/978-3-642-16014-1_19
- 39 Jiang, L., Guli-Jiapaer, Bao, A., Guo, H., & Ndayisaba, F. (2017). Vegetation dynamics and responses
40 to climate change and human activities in Central Asia. *Science of The Total Environment*, 599–
41 600, 967–980. <https://doi.org/10.1016/J.SCITOTENV.2017.05.012>
- 42 Jiang, Z., Lian, Y., & Qin, X. (2014). Rocky desertification in Southwest China: Impacts, causes, and
43 restoration. *Earth-Science Reviews*, 132, 1–12.
44 <https://doi.org/10.1016/J.EARSCIREV.2014.01.005>
- 45 Jickells, T. D., An, Z. S., Andersen, K. K., Baker, A. R., Bergametti, G., Brooks, N., ... Torres, R.
46 (2005). Global iron connections between desert dust, ocean biogeochemistry, and climate.
47 *Science (New York, N.Y.)*, 308(5718), 67–71. <https://doi.org/10.1126/science.1105959>

- 1 Jilali, A. (2014). Impact of climate change on the Figuig aquifer using a numerical model: Oasis of
2 Eastern Morocco. *Journal of Biology and Earth Sciences*, 4(1), 16–24. Retrieved from
3 https://www.researchgate.net/publication/267579009_Impact_of_climate_change_on_the_Figuig_aquifer_using_a_numerical_model_Oasis_of_Eastern_Morocco
4
- 5 Jiménez Cisneros, B. E., Oki, T., Arnell, N. W., Benito, G., Cogley, J. G., Döll, P., ... Mwakalila, S.
6 S. (2014). Freshwater resources. In C. B. Field, V. R. Barros, D. J. Dokken, K. J. Mach, M. D.
7 Mastrandrea, T. E. Bilir, ... L.L.White (Eds.), *Climate Change 2014: Impacts, Adaptation, and*
8 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the*
9 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change* (pp. 229–269).
10 Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA.
- 11 Jin, Z., Dong, Y., Wang, Y., Wei, X., Wang, Y., Cui, B., & Zhou, W. (2014). Natural vegetation
12 restoration is more beneficial to soil surface organic and inorganic carbon sequestration than tree
13 plantation on the Loess Plateau of China. *Science of The Total Environment*, 485–486, 615–623.
14 <https://doi.org/10.1016/J.SCITOTENV.2014.03.105>
- 15 Jish Prakash, P., Stenchikov, G., Kalenderski, S., Osipov, S., & Bangalath, H. (2015). The impact of
16 dust storms on the Arabian Peninsula and the Red Sea. *Atmospheric Chemistry and Physics*,
17 15(1), 199–222. <https://doi.org/10.5194/acp-15-199-2015>
- 18 Johannsen, I., Hengst, J., Goll, A., Höllermann, B., Diekkrüger, B., Johannsen, I. M., ... Diekkrüger,
19 B. (2016). Future of Water Supply and Demand in the Middle Drâa Valley, Morocco, under
20 Climate and Land Use Change. *Water*, 8(8), 313. <https://doi.org/10.3390/w8080313>
- 21 Johanson, C. M., Fu, Q., Johanson, C. M., & Fu, Q. (2009). Hadley Cell Widening: Model
22 Simulations versus Observations. *Journal of Climate*, 22(10), 2713–2725.
23 <https://doi.org/10.1175/2008JCLI2620.1>
- 24 Johnstone, S., & Mazo, J. (2011). Global Warming and the Arab Spring. *Survival*, 53(2), 11–17.
25 <https://doi.org/10.1080/00396338.2011.571006>
- 26 Jolly, I. D., Mcewan, K. L., & Holland, K. L. (2008). A review of groundwater – surface water
27 interactions in arid / semi-arid wetlands and the consequences of salinity for wetland ecology,
28 58(January), 43–58. <https://doi.org/10.1002/eco>
- 29 Jolly, W. M., Cochrane, M. A., Freeborn, P. H., Holden, Z. A., Brown, T. J., Williamson, G. J., &
30 Bowman, D. M. J. S. (2015). Climate-induced variations in global wildfire danger from 1979 to
31 2013. *Nature Communications*, 6(May), 1–11. <https://doi.org/10.1038/ncomms8537>
- 32 Jones, M. D., Abu-Jaber, N., AlShdaifat, A., Baird, D., Cook, B. I., Cuthbert, M. O., ... Weeks, L.
33 (2019). 20,000 years of societal vulnerability and adaptation to climate change in southwest
34 Asia. *Wiley Interdisciplinary Reviews: Water*, 6(2), e1330. <https://doi.org/10.1002/wat2.1330>
- 35 Joy, K. ., Paranjape, S., Kirankumar, A. ., Lele, R., & Adagale, R. (2004). *Watershed development*
36 *review: issues and prospects*. Bangalore, India.
- 37 Joy, K. J., Shah, A., Paranjape, S., Badiger, S., & Lélé, S. (2006). Issues in restructuring. *Economic*
38 *and Political Weekly*, 2994–2996.
- 39 Júnior, W. S. F., Santoro, F. R., Vandebroek, I., & Albuquerque, U. P. (2016). Urbanization,
40 Modernization, and Nature Knowledge. In *Introduction to Ethnobiology* (pp. 251–256). Cham:
41 Springer International Publishing. https://doi.org/10.1007/978-3-319-28155-1_37
- 42 Jury, M. R. (2016). Climate influences on upper Limpopo River flow. *Water SA*, 42(1), 63.
43 <https://doi.org/10.4314/wsa.v42i1.08>
- 44 Kadirbeyoglu, Z., & Özertan, G. (2015). Power in the Governance of Common-Pool Resources: A
45 comparative analysis of irrigation management decentralization in Turkey. *Environmental*
46 *Policy and Governance*, 25(3), 157–171. <https://doi.org/10.1002/eet.1673>
- 47 Kahinda, J. M., Meissner, R., & Engelbrecht, F. A. (2016). Implementing Integrated Catchment
48 Management in the upper Limpopo River basin: A situational assessment. *Physics and*

- 1 *Chemistry of the Earth*, 93(January 2012), 104–118. <https://doi.org/10.1016/j.pce.2015.10.003>
- 2 Kaijser, A., & Kronsell, A. (2014). Climate change through the lens of intersectionality.
3 *Environmental Politics*, 23(3), 417–433. <https://doi.org/10.1080/09644016.2013.835203>
- 4 Kamali, B., Abbaspour, K. C., Wehrli, B., & Yang, H. (2018). Drought vulnerability assessment of
5 maize in Sub-Saharan Africa: Insights from physical and social perspectives. *Global and*
6 *Planetary Change*, 162(January), 266–274. <https://doi.org/10.1016/j.gloplacha.2018.01.011>
- 7 Kantarcı, M. D., Özel, H. B., Ertek, M. and Kırdar, E. (2011). An assessment on the adaptation of 6
8 tree species to steppe habitat during Konya-Karapınar sand-dune afforestations. *Journal of*
9 *Bartın Forestry Faculty*, 13(19), 107–127. Retrieved from
10 <https://www.cabdirect.org/cabdirect/abstract/20113318864>
- 11 Kapović Solomun, M., Barger, N., Cerda, A., Keesstra, S., & Marković, M. (2018). Assessing land
12 condition as a first step to achieving land degradation neutrality: A case study of the Republic of
13 Srpska. *Environmental Science and Policy*, 90, 19–27.
14 <https://doi.org/10.1016/j.envsci.2018.09.014>
- 15 Kaptué, A. T., Prihodko, L., & Hanan, N. P. (2015). On greening and degradation in Sahelian
16 watersheds. *Proceedings of the National Academy of Sciences of the United States of America*,
17 112(39), 12133–8. <https://doi.org/10.1073/pnas.1509645112>
- 18 Karami Dehkordi, M., Kohestani, H., Yadavar, H., Roshandel, R., & Karbasioun, M. (2017).
19 Implementing conceptual model using renewable energies in rural area of Iran. *Information*
20 *Processing in Agriculture*, 4(3), 228–240. <https://doi.org/10.1016/J.INPA.2017.02.003>
- 21 Karatayev, M., Hall, S., Kalyuzhnova, Y., & Clarke, M. L. (2016). Renewable energy technology
22 uptake in Kazakhstan: Policy drivers and barriers in a transitional economy. *Renewable and*
23 *Sustainable Energy Reviews*, 66, 120–136. <https://doi.org/10.1016/J.RSER.2016.07.057>
- 24 Karmaoui, A., Ifaadassan, I., Babqiqi, A., Messouli, M., & Khebiza, Y. M. (2016). Analysis of the
25 Water Supply-demand Relationship in the Middle Draa Valley, Morocco, under Climate Change
26 and Socio-economic Scenarios. *Journal of Scientific Research & Reports*, 9(4), 1–10. Retrieved
27 from [https://www.academia.edu/35130199/Analysis_of_the_Water_Supply-](https://www.academia.edu/35130199/Analysis_of_the_Water_Supply-demand_Relationship_in_the_Middle_Draa_Valley_Morocco_under_Climate_Change_and_Socio-economic_Scenarios)
28 [demand_Relationship_in_the_Middle_Draa_Valley_Morocco_under_Climate_Change_and_Soc](https://www.academia.edu/35130199/Analysis_of_the_Water_Supply-demand_Relationship_in_the_Middle_Draa_Valley_Morocco_under_Climate_Change_and_Socio-economic_Scenarios)
29 [io-economic_Scenarios](https://www.academia.edu/35130199/Analysis_of_the_Water_Supply-demand_Relationship_in_the_Middle_Draa_Valley_Morocco_under_Climate_Change_and_Socio-economic_Scenarios)
- 30 Karsenty, A., & Ongolo, S. (2012). Can “fragile states” decide to reduce their deforestation? The
31 inappropriate use of the theory of incentives with respect to the REDD mechanism. *Forest*
32 *Policy and Economics*, 18, 38–45. <https://doi.org/10.1016/J.FORPOL.2011.05.006>
- 33 Kashima, S., Yorifuji, T., Bae, S., Honda, Y., Lim, Y.-H., & Hong, Y.-C. (2016). Asian dust effect on
34 cause-specific mortality in five cities across South Korea and Japan. *Atmospheric Environment*,
35 128, 20–27. <https://doi.org/10.1016/J.ATMOENV.2015.12.063>
- 36 Kaspari, S., Painter, T. H., Gysel, M., Skiles, S. M., & Schwikowski, M. (2014). Seasonal and
37 elevational variations of black carbon and dust in snow and ice in the Solu-Khumbu, Nepal and
38 estimated radiative forcings. *Atmospheric Chemistry and Physics*, 14(15), 8089–8103.
39 <https://doi.org/10.5194/acp-14-8089-2014>
- 40 Kassahun, A., Snyman, H. A., & Smit, G. N. (2008). Impact of rangeland degradation on the pastoral
41 production systems, livelihoods and perceptions of the Somali pastoralists in Eastern Ethiopia.
42 *Journal of Arid Environments*, 72(7), 1265–1281.
43 <https://doi.org/10.1016/J.JARIDENV.2008.01.002>
- 44 Kassas, M. (1995). Desertification: A general review. *Journal of Arid Environments*, 30(2), 115–128.
45 [https://doi.org/10.1016/S0140-1963\(05\)80063-1](https://doi.org/10.1016/S0140-1963(05)80063-1)
- 46 Kassie, M., Jaleta, M., Shiferaw, B., Mmbando, F., & Mekuria, M. (2013). Adoption of interrelated
47 sustainable agricultural practices in smallholder systems: Evidence from rural Tanzania.
48 *Technological Forecasting and Social Change*, 80(3), 525–540.

- 1 <https://doi.org/10.1016/J.TECHFORE.2012.08.007>
- 2 Kassie, M., Teklewold, H., Jaleta, M., Marenya, P., & Erenstein, O. (2015). Understanding the
3 adoption of a portfolio of sustainable intensification practices in eastern and southern Africa.
4 *Land Use Policy*, 42, 400–411. <https://doi.org/10.1016/J.LANDUSEPOL.2014.08.016>
- 5 Kath, J., Mushtaq, S., Henry, R., Adeyinka, A. A., Stone, R., Marcussen, T., & Kouadio, L. (2019).
6 Spatial variability in regional scale drought index insurance viability across Australia’s wheat
7 growing regions. *Climate Risk Management*, 24, 13–29.
8 <https://doi.org/10.1016/J.CRM.2019.04.002>
- 9 Katikiro, R. E. (2016). Prospects for the Uptake of Renewable Energy Technologies in Rural
10 Tanzania. *Energy Procedia*, 93, 229–233. <https://doi.org/10.1016/J.EGYPRO.2016.07.175>
- 11 Kaufman, Y. J., Koren, I., Remer, L. A., Tanré, D., Ginoux, P., & Fan, S. (2005). Dust transport and
12 deposition observed from the Terra-Moderate Resolution Imaging Spectroradiometer (MODIS)
13 spacecraft over the Atlantic Ocean. *Journal of Geophysical Research*, 110(D10), D10S12.
14 <https://doi.org/10.1029/2003JD004436>
- 15 Kaufman, Y. J., Tanré, D., & Boucher, O. (2002). A satellite view of aerosols in the climate system.
16 *Nature*, 419(6903), 215–223. <https://doi.org/10.1038/nature01091>
- 17 Kazbekov, J., Abdullaev, I., Manthritilake, H., Qureshi, A., & Jumaboev, K. (2009). Evaluating
18 planning and delivery performance of Water User Associations (WUAs) in Osh Province,
19 Kyrgyzstan. *Agricultural Water Management*, 96(8), 1259–1267.
20 <https://doi.org/10.1016/J.AGWAT.2009.04.002>
- 21 Keeley, J. E., & Brennan, T. J. (2012). Fire-driven alien invasion in a fire-adapted ecosystem.
22 *Oecologia*, 169(4), 1043–1052. <https://doi.org/10.1007/s00442-012-2253-8>
- 23 Kéfi, S., Alados, C. L., Chaves, R. C. G., Pueyo, Y., & Rietkerk, M. (2010). Is the patch size
24 distribution of vegetation a suitable indicator of desertification processes? Comment. *Ecology*,
25 91(12), 3739–3742. <https://doi.org/10.1890/09-1915.1>
- 26 Keiichiro, H., Otsuki, T., & Wilson, J. S. (2015). Food Safety Standards and International Trade: The
27 Impact on Developing Countries’ Export Performance. In *Food Safety, Market Organization,*
28 *Trade and Development* (pp. 151–166). Cham: Springer International Publishing.
29 https://doi.org/10.1007/978-3-319-15227-1_8
- 30 Kellogg, C. A., Griffin, D. W., Garrison, V. H., Peak, K. K., Royall, N., Smith, R. R., & Shinn, E. A.
31 (2004). Characterization of aerosolized bacteria and fungi from desert dust events in Mali,
32 West Africa. *Africa*, (Figure 1), 99–110.
- 33 Kerr, J. (2007). Watershed Management: Lessons from Common Property Theory. *International*
34 *Journal of the Commons*, 1, 89–109. Retrieved from
35 [https://www.jstor.org/stable/pdf/26522983.pdf?refreqid=excelsior%3Acdd4d8aabb9a055e12](https://www.jstor.org/stable/pdf/26522983.pdf?refreqid=excelsior%3Acdd4d8aabb9a055e1219f49df7b73)
36 [19f49df7b73](https://www.jstor.org/stable/pdf/26522983.pdf?refreqid=excelsior%3Acdd4d8aabb9a055e1219f49df7b73)
- 37 Kerr, J., Pangare, G., & Pangare, V. (2002). *Watershed development projects in India: an evaluation.*
38 Retrieved from
39 [https://books.google.com/books?hl=en&lr=&id=3QlsCU6_o5UC&oi=fnd&pg=PR4&dq=Kerr,+](https://books.google.com/books?hl=en&lr=&id=3QlsCU6_o5UC&oi=fnd&pg=PR4&dq=Kerr,+J.+M.,+G.+Pangare,+and+V.+Pangare,+2002:+Watershed+development+projects+in+India:+an+evaluation.+Intl+Food+Policy+Res+Inst.,&ots=Aem0Fd0yP-&sig=Gq1_zR70Gm99-jN4589rnNsbxG8)
40 [J.+M.,+G.+Pangare,+and+V.+Pangare,+2002:+Watershed+development+projects+in+India:+an](https://books.google.com/books?hl=en&lr=&id=3QlsCU6_o5UC&oi=fnd&pg=PR4&dq=Kerr,+J.+M.,+G.+Pangare,+and+V.+Pangare,+2002:+Watershed+development+projects+in+India:+an+evaluation.+Intl+Food+Policy+Res+Inst.,&ots=Aem0Fd0yP-&sig=Gq1_zR70Gm99-jN4589rnNsbxG8)
41 [+evaluation.+Intl+Food+Policy+Res+Inst.,&ots=Aem0Fd0yP-&sig=Gq1_zR70Gm99-](https://books.google.com/books?hl=en&lr=&id=3QlsCU6_o5UC&oi=fnd&pg=PR4&dq=Kerr,+J.+M.,+G.+Pangare,+and+V.+Pangare,+2002:+Watershed+development+projects+in+India:+an+evaluation.+Intl+Food+Policy+Res+Inst.,&ots=Aem0Fd0yP-&sig=Gq1_zR70Gm99-jN4589rnNsbxG8)
42 [jN4589rnNsbxG8](https://books.google.com/books?hl=en&lr=&id=3QlsCU6_o5UC&oi=fnd&pg=PR4&dq=Kerr,+J.+M.,+G.+Pangare,+and+V.+Pangare,+2002:+Watershed+development+projects+in+India:+an+evaluation.+Intl+Food+Policy+Res+Inst.,&ots=Aem0Fd0yP-&sig=Gq1_zR70Gm99-jN4589rnNsbxG8)
- 43 Kertis, C. A. (2003). Soil Erosion on Cropland in The United States: Status and Trends for 1982-
44 2003. Retrieved May 26, 2018, from
45 [http://citeseerx.ist.psu.edu/viewdoc/versions;jsessionid=D92B50C8C236BD96C4F1D967FA7E](http://citeseerx.ist.psu.edu/viewdoc/versions;jsessionid=D92B50C8C236BD96C4F1D967FA7E0763?doi=10.1.1.613.529)
46 [0763?doi=10.1.1.613.529](http://citeseerx.ist.psu.edu/viewdoc/versions;jsessionid=D92B50C8C236BD96C4F1D967FA7E0763?doi=10.1.1.613.529)
- 47 Kgope, B. S., Bond, W. J., & Midgley, G. F. (2010). Growth responses of African savanna trees
48 implicate atmospheric [CO₂] as a driver of past and current changes in savanna tree cover.

- 1 *Austral Ecology*, 35, 451–463.
- 2 Khamzina, A., Lamers, J. P. A., & Martius, C. (2016). Above- and belowground litter stocks and
3 decay at a multi-species afforestation site on arid, saline soil. *Nutrient Cycling in*
4 *Agroecosystems*, 104(2), 187–199. <https://doi.org/10.1007/s10705-016-9766-1>
- 5 Khamzina, A., Lamers, J. P. A., & Vlek, P. L. G. (2009). Nitrogen fixation by *Elaeagnus angustifolia*
6 in the reclamation of degraded croplands of Central Asia. *Tree Physiology*, 29(6), 799–808.
7 <https://doi.org/10.1093/treephys/tpp017>
- 8 Khan, M. A., Qureshi, R. A., Gillani, S. A., Ghufuran, M. A., Batool, A., & Sultana, K. N. (2010).
9 Invasive species of federal capital area Islamabad, Pakistan. *Pakistan Journal of Botany*, 42(3),
10 1529–1534.
- 11 Khan, Z. R., Midega, C. A. O., Pittchar, J. O., Murage, A. W., Birkett, M. A., Bruce, T. J. A., &
12 Pickett, J. A. (2014). Achieving food security for one million sub-Saharan African poor through
13 push-pull innovation by 2020. *Philosophical Transactions of the Royal Society B: Biological*
14 *Sciences*, 369(1639), 20120284–20120284. <https://doi.org/10.1098/rstb.2012.0284>
- 15 Khishigbayar, J., Fernández-Giménez, M. E., Angerer, J. P., Reid, R. S., Chantsallkham, J.,
16 Baasandorj, Y., & Zumberelmaa, D. (2015). Mongolian rangelands at a tipping point? Biomass
17 and cover are stable but composition shifts and richness declines after 20years of grazing and
18 increasing temperatures. *Journal of Arid Environments*, 115, 100–112.
19 <https://doi.org/10.1016/j.jaridenv.2015.01.007>
- 20 Kidron, G. J., & Gutschick, V. P. (2017). Temperature rise may explain grass depletion in the
21 Chihuahuan Desert. *Ecohydrology*, 10(4), e1849. <https://doi.org/10.1002/eco.1849>
- 22 Kienzler, K. M., Lamers, J. P. A., McDonald, A., Mirzabaev, A., Ibragimov, N., Egamberdiev, O., ...
23 Akramkhanov, A. (2012). Conservation agriculture in Central Asia-What do we know and
24 where do we go from here? *Field Crops Research*, 132. <https://doi.org/10.1016/j.fcr.2011.12.008>
- 25 Kihui, E. (2016). *Pastoral Practices, Economics, and Institutions of Sustainable Rangeland*
26 *Management in Kenya*. University of Bonn. Retrieved from [http://hss.ulb.uni-](http://hss.ulb.uni-bonn.de/2016/4315/4315.htm)
27 [bonn.de/2016/4315/4315.htm](http://hss.ulb.uni-bonn.de/2016/4315/4315.htm)
- 28 Kihui, E. N. (2016). Basic capability effect: Collective management of pastoral resources in
29 southwestern Kenya. *Ecological Economics*, 123, 23–34.
30 <https://doi.org/10.1016/J.ECOLECON.2016.01.003>
- 31 Kim, S. E., Honda, Y., Hashizume, M., Kan, H., Lim, Y.-H., Lee, H., ... Kim, H. (2017). Seasonal
32 analysis of the short-term effects of air pollution on daily mortality in Northeast Asia. *Science of*
33 *The Total Environment*, 576, 850–857. <https://doi.org/10.1016/J.SCITOTENV.2016.10.036>
- 34 King, C., & Thomas, D. S. G. (2014). Monitoring environmental change and degradation in the
35 irrigated oases of the Northern Sahara. *Journal of Arid Environments*, 103, 36–45.
36 <https://doi.org/10.1016/J.JARIDENV.2013.12.009>
- 37 Kingdom of Saudi Arabia. (2016). *National transformation program 2020. Saudi Vision 2030*.
- 38 Kiptot, E., Franzel, S., & Degrande, A. (2014). Gender, agroforestry and food security in Africa.
39 *Current Opinion in Environmental Sustainability*, 6, 104–109.
40 <https://doi.org/10.1016/J.COSUST.2013.10.019>
- 41 Kirui, O. K. (2016). *Economics of Land Degradation, Sustainable Land Management and Poverty in*
42 *Eastern Africa - The Extent, Drivers, Costs and Impacts*. University of Bonn. Retrieved from
43 <http://hss.ulb.uni-bonn.de/2016/4355/4355.pdf>
- 44 Klein Goldewijk, K., Beusen, A., Doelman, J., & Stehfest, E. (2017). Anthropogenic land use
45 estimates for the Holocene – HYDE 3.2. *Earth System Science Data*, 9(2), 927–953.
46 <https://doi.org/10.5194/essd-9-927-2017>
- 47 Klein, R., Midgley, J. T., Preston, G. F., L., B., Alam, M., Berkhout, F. G. H., ... Thomas, A. (2015).

- 1 Adaptation opportunities, constraints, and limits. *Climate Change 2014 Impacts, Adaptation and*
2 *Vulnerability: Part A: Global and Sectoral Aspects*, 899–944.
3 <https://doi.org/10.1017/CBO9781107415379.021>
- 4 Klik, A., & Eitzinger, J. (2010). Impact of climate change on soil erosion and the efficiency of soil
5 conservation practices in Austria. *Journal of Agricultural Science*, 148(5), 529–541.
6 <https://doi.org/10.1017/S0021859610000158>
- 7 Klingmüller, K., Pozzer, A., Metzger, S., Stenchikov, G. L., & Lelieveld, J. (2016). Aerosol optical
8 depth trend over the Middle East. *Atmospheric Chemistry and Physics*, 16(8), 5063–5073.
9 <https://doi.org/10.5194/acp-16-5063-2016>
- 10 Kloos, J., & Renaud, F. G. (2016). Overview of Ecosystem-Based Approaches to Drought Risk
11 Reduction Targeting Small-Scale Farmers in Sub-Saharan Africa (pp. 199–226). Springer,
12 Cham. https://doi.org/10.1007/978-3-319-43633-3_9
- 13 Knight, J. R., Folland, C. K., & Scaife, A. A. (2006). Climate impacts of the Atlantic Multidecadal
14 Oscillation. *Geophysical Research Letters*, 33(17), L17706.
15 <https://doi.org/10.1029/2006GL026242>
- 16 Kniveton, D., Smith, C., & Wood, S. (2011). Agent-based model simulations of future changes in
17 migration flows for Burkina Faso. *Global Environmental Change*, 21, S34–S40.
18 <https://doi.org/10.1016/J.GLOENVCHA.2011.09.006>
- 19 Knörzer, H., Graeff-Hönninger, S., Guo, B., Wang, P., & Claupein, W. (2009). The Rediscovery of
20 Intercropping in China: A Traditional Cropping System for Future Chinese Agriculture – A
21 Review. In *Climate Change, Intercropping, Pest Control and Beneficial Microorganisms* (pp.
22 13–44). Dordrecht: Springer Netherlands. https://doi.org/10.1007/978-90-481-2716-0_3
- 23 Knox, J., Hess, T., Daccache, A., & Wheeler, T. (2012). Climate change impacts on crop productivity
24 in Africa and South Asia. *Environmental Research Letters*, 7(3), 034032.
25 <https://doi.org/10.1088/1748-9326/7/3/034032>
- 26 Knutson, K. C., Pyke, D. A., Wirth, T. A., Arkle, R. S., Pilliod, D. S., Brooks, M. L., ... Grace, J. B.
27 (2014). Long-term effects of seeding after wildfire on vegetation in Great Basin shrubland
28 ecosystems. *Journal of Applied Ecology*, 51(5), 1414–1424. [https://doi.org/10.1111/1365-](https://doi.org/10.1111/1365-2664.12309)
29 2664.12309
- 30 Kodirekkala, K. R. (2017). Internal and External Factors Affecting Loss of Traditional Knowledge:
31 Evidence from a Horticultural Society in South India. *Journal of Anthropological Research*,
32 73(1), 22–42. <https://doi.org/10.1086/690524>
- 33 Kok, J. F., Ward, D. S., Mahowald, N. M., & Evan, A. T. (2018). Global and regional importance of
34 the direct dust-climate feedback. *Nature Communications*, 9(1), 241.
35 <https://doi.org/10.1038/s41467-017-02620-y>
- 36 Kok, M., Lüdeke, M., Lucas, P., Sterzel, T., Walther, C., Janssen, P., ... de Soysa, I. (2016). A new
37 method for analysing socio-ecological patterns of vulnerability. *Regional Environmental*
38 *Change*, 16(1), 229–243. <https://doi.org/10.1007/s10113-014-0746-1>
- 39 Komala, P., & Prasad, G. (2016). Biomass: A key source of energy in rural households of
40 Chamarajanagar district. *Pelagia Research Library Advances in Applied Science Research*, 7(1),
41 85–89. Retrieved from www.pelagiaresearchlibrary.com
- 42 Konare, A., Zakey, A. S., Solmon, F., Giorgi, F., Rauscher, S., Ibrah, S., & Bi, X. (2008). A regional
43 climate modeling study of the effect of desert dust on the West African monsoon. *Journal of*
44 *Geophysical Research*, 113(D12), D12206. <https://doi.org/10.1029/2007JD009322>
- 45 Konikow, L. F. (2011). Contribution of global groundwater depletion since 1900 to sea-level rise.
46 *Geophysical Research Letters*, 38(17), 1–5. <https://doi.org/10.1029/2011GL048604>
- 47 Kouba, Y., Gartzia, M., El Aich, A., & Alados, C. L. (2018). Deserts do not advance, they are created:
48 Land degradation and desertification in semiarid environments in the Middle Atlas, Morocco.

- 1 *Journal of Arid Environments*, 158, 1–8. <https://doi.org/10.1016/J.JARIDENV.2018.07.002>
- 2 Koubi, V., Spilker, G., Schaffer, L., & Bernauer, T. (2016). Environmental Stressors and Migration:
3 Evidence from Vietnam. *World Development*, 79, 197–210.
4 <https://doi.org/10.1016/J.WORLDDEV.2015.11.016>
- 5 Koutroulis, A. G. (2019). Dryland changes under different levels of global warming. *Science of The*
6 *Total Environment*, 655, 482–511. <https://doi.org/10.1016/J.SCITOTENV.2018.11.215>
- 7 Koyro, H.-W., Hussain, T., Huchzermeyer, B., & Khan, M. A. (2013). Photosynthetic and growth
8 responses of a perennial halophytic grass *Panicum turgidum* to increasing NaCl concentrations.
9 *Environmental and Experimental Botany*, 91, 22–29.
10 <https://doi.org/10.1016/J.ENVEXPBOT.2013.02.007>
- 11 Krätli, S., & Schareika, N. (2010). Living Off Uncertainty: The Intelligent Animal Production of
12 Dryland Pastoralists. *The European Journal of Development Research*, 22(5), 605–622.
13 <https://doi.org/10.1057/ejdr.2010.41>
- 14 Kubik, Z., & Maurel, M. (2016). Weather Shocks, Agricultural Production and Migration: Evidence
15 from Tanzania. *The Journal of Development Studies*, 52(5), 665–680.
16 <https://doi.org/10.1080/00220388.2015.1107049>
- 17 Kundu, A., Patel, N. R., Saha, S. K., & Dutta, D. (2017). Desertification in western Rajasthan (India):
18 an assessment using remote sensing derived rain-use efficiency and residual trend methods.
19 *Natural Hazards*, 86(1), 297–313. <https://doi.org/10.1007/s11069-016-2689-y>
- 20 Kust, G. (1999). *Desertification: principles of ecological assessment and mapping*. Moscow: MSU-
21 RAS Institute of Soil Science.
- 22 Kust, G., Mott, J., Jain, N., Sampath, T., & Armstrong, A. (2014). SLM Oriented Projects in
23 Tajikistan: Experience and Lessons Learned. *Planet@Risk*, 2(1). Retrieved from [https://planet-](https://planet-risk.org/index.php/pr/article/view/31/147)
24 [risk.org/index.php/pr/article/view/31/147](https://planet-risk.org/index.php/pr/article/view/31/147)
- 25 Kust, G. S., Andreeva, O. V., & Dobrynin, D. V. (2011). Desertification assessment and mapping in
26 the Russian Federation. *Arid Ecosystems*, 1(1), 14–28.
27 <https://doi.org/10.1134/S2079096111010057>
- 28 Kusunose, Y., & Lybbert, T. J. (2014). Coping with Drought by Adjusting Land Tenancy Contracts:
29 A Model and Evidence from Rural Morocco. *World Development*, 61, 114–126.
30 <https://doi.org/10.1016/J.WORLDDEV.2014.04.006>
- 31 Lahlaoui, H., Rhinane, H., Hilali, A., Lahssini, S., Moukrim, S., Lahlaoui, H., ... Moukrim, S. (2017).
32 Desertification Assessment Using MEDALUS Model in Watershed Oued El Maleh, Morocco.
33 *Geosciences*, 7(3), 50. <https://doi.org/10.3390/geosciences7030050>
- 34 Lal, R., Stewart, B. A. (2012). *Advances in Soil Science: Soil Water and Agronomic Productivity*. (B.
35 A. Lal, R., Stewart, Ed.). Boca Raton, Florida: CRC Press.
- 36 Lal, R. (1997). Degradation and resilience of soils. *Philosophical Transactions of the Royal Society of*
37 *London. Series B: Biological Sciences*, 352(1356), 997–1010.
38 <https://doi.org/10.1098/rstb.1997.0078>
- 39 Lal, R. (2003). Carbon Sequestration in Dryland Ecosystems. *Environmental Management*, 33(April),
40 528–544. <https://doi.org/10.1007/s00267-003-9110-9>
- 41 Lal, R. (2004). Carbon Sequestration in Dryland Ecosystems. *Environmental Management*, 33(4),
42 528–544. <https://doi.org/10.1007/s00267-003-9110-9>
- 43 Lal, R. (2009). Sequestering carbon in soils of arid ecosystems. *Land Degradation & Development*,
44 20(4), 441–454. <https://doi.org/10.1002/ldr.934>
- 45 Lal, R. (2015). The soil–peace nexus: our common future. *Soil Science and Plant Nutrition*, 61(4),
46 566–578. <https://doi.org/10.1080/00380768.2015.1065166>

- 1 Lal, R. (2018). Managing agricultural soils of Pakistan for food and climate. *Soil and Environment*,
2 37(1), 1–10.
- 3 Lal, R. (ed. . (2016). *Encyclopedia of soil science*. CRC Press. Retrieved from
4 [https://www.crcpress.com/Encyclopedia-of-Soil-Science-Third-](https://www.crcpress.com/Encyclopedia-of-Soil-Science-Third-Edition/Lal/p/book/9781498738903)
5 [Edition/Lal/p/book/9781498738903](https://www.crcpress.com/Encyclopedia-of-Soil-Science-Third-Edition/Lal/p/book/9781498738903)
- 6 Lam, D. K., Rimmel, T. K., Drezner, T. D., Lam, D. K., Rimmel, T. K., & Drezner, T. D. (2011).
7 Tracking Desertification in California Using Remote Sensing: A Sand Dune Encroachment
8 Approach. *Remote Sensing*, 3(1), 1–13. <https://doi.org/10.3390/rs3010001>
- 9 Lambin, E. F. (2012). Global land availability: Malthus versus Ricardo. *Global Food Security*, 1(2),
10 83–87. <https://doi.org/10.1016/J.GFS.2012.11.002>
- 11 Lambin, E. F., Gibbs, H. K., Ferreira, L., Grau, R., Mayaux, P., Meyfroidt, P., ... Munger, J. (2013).
12 Estimating the world's potentially available cropland using a bottom-up approach. *Global*
13 *Environmental Change*, 23(5), 892–901. <https://doi.org/10.1016/J.GLOENVCHA.2013.05.005>
- 14 Lambin, E. F., & Meyfroidt, P. (2011). Global land use change, economic globalization, and the
15 looming land scarcity. *Proceedings of the National Academy of Sciences of the United States of*
16 *America*, 108(9), 3465–72. <https://doi.org/10.1073/pnas.1100480108>
- 17 Lambin, E. F., Meyfroidt, P., Rueda, X., Blackman, A., Börner, J., Cerutti, P. O., ... Wunder, S.
18 (2014). Effectiveness and synergies of policy instruments for land use governance in tropical
19 regions. *Global Environmental Change*, 28, 129–140.
20 <https://doi.org/10.1016/J.GLOENVCHA.2014.06.007>
- 21 Lamchin, M., Lee, J.-Y., Lee, W.-K., Lee, E. J., Kim, M., Lim, C.-H., ... Kim, S.-R. (2016).
22 Assessment of land cover change and desertification using remote sensing technology in a local
23 region of Mongolia. *Advances in Space Research*, 57(1), 64–77.
24 <https://doi.org/10.1016/J.ASR.2015.10.006>
- 25 Lamontagne, S., Cook, P. G., O'Grady, A., & Eamus, D. (2005). Groundwater use by vegetation in a
26 tropical savanna riparian zone (Daly River, Australia). *Journal of Hydrology*, 310(1–4), 280–
27 293. <https://doi.org/10.1016/J.JHYDROL.2005.01.009>
- 28 Lands, D., & Nations, U. (2011). Publisher : Arab Center for the Studies of Arid Zones and Dry Lands
29 Responsible / Contact : Design :
- 30 Larrucea, E. S., & Brussard, P. F. (2008). Habitat selection and current distribution of the pygmy
31 rabbit in Nevada and California, USA. *Journal of Mammalogy*, 89(3), 691–699.
32 <https://doi.org/10.1644/07-MAMM-A-199R.1>
- 33 Lau, K. M., Kim, K. M., Sud, Y. C., & Walker, G. K. (2009). A GCM study of the response of the
34 atmospheric water cycle of West Africa and the Atlantic to Saharan dust radiative forcing.
35 *Annales Geophysicae*, 27(10), 4023–4037. <https://doi.org/10.5194/angeo-27-4023-2009>
- 36 Lavauden, L. (1927). *Les forêts du Sahara*. Paris, France: Berger-Levrault.
- 37 Lavee, H., Imeson, A. C., & Sarah, P. (1998). The impact of climate change on geo-morphology and
38 desertification along a mediterranean arid transect. *Land Degradation and Development*,
39 9(SEPTEMBER), 407–422. [https://doi.org/10.1002/\(SICI\)1099-](https://doi.org/10.1002/(SICI)1099-145X(199809/10)9:5<407::AID-LDR302>3.0.CO;2-6)
40 [145X\(199809/10\)9:5<407::AID-LDR302>3.0.CO;2-6](https://doi.org/10.1002/(SICI)1099-145X(199809/10)9:5<407::AID-LDR302>3.0.CO;2-6)
- 41 Lawrence, P. G., Maxwell, B. D., Rew, L. J., Ellis, C., & Bekkerman, A. (2018). Vulnerability of
42 dryland agricultural regimes to economic and climatic change. *Ecology and Society*, 23(1),
43 art34. <https://doi.org/10.5751/ES-09983-230134>
- 44 Lawry, S., Samii, C., Hall, R., Leopold, A., Hornby, D., & Mtero, F. (2017). The impact of land
45 property rights interventions on investment and agricultural productivity in developing
46 countries: a systematic review. *Journal of Development Effectiveness*, 9(1), 61–81.
47 <https://doi.org/10.1080/19439342.2016.1160947>

- 1 Lázaro, R., Cantón, Y., Solé-Benet, A., Bevan, J., Alexander, R., Sancho, L. G., & Puigdefábregas, J.
2 (2008). The influence of competition between lichen colonization and erosion on the evolution
3 of soil surfaces in the Tabernas badlands (SE Spain) and its landscape effects. *Geomorphology*,
4 *102*(2), 252–266. <https://doi.org/10.1016/J.GEOMORPH.2008.05.005>
- 5 Le Houérou, H. N. (1996). Climate change, drought and desertification. *Journal of Arid*
6 *Environments*, *34*(2), 133–185. <https://doi.org/10.1006/jare.1996.0099>
- 7 Le Maitre, D. C., Gaertner, M., Marchante, E., Ens, E.-J., Holmes, P. M., Pauchard, A., ...
8 Richardson, D. M. (2011). Impacts of invasive Australian acacias: implications for management
9 and restoration. *Diversity and Distributions*, *17*(5), 1015–1029. [https://doi.org/10.1111/j.1472-](https://doi.org/10.1111/j.1472-4642.2011.00816.x)
10 [4642.2011.00816.x](https://doi.org/10.1111/j.1472-4642.2011.00816.x)
- 11 Le Maitre, D. C., Gush, M. B., & Dzikiti, S. (2015). Impacts of invading alien plant species on water
12 flows at stand and catchment scales. *AoB Plants*, *7*, plv043.
13 <https://doi.org/10.1093/aobpla/plv043>
- 14 Le, Q. B., Nkonya, E., & Mirzabaev, A. (2016). Biomass Productivity-Based Mapping of Global
15 Land Degradation Hotspots. In *Economics of Land Degradation and Improvement – A Global*
16 *Assessment for Sustainable Development* (pp. 55–84). Cham: Springer International Publishing.
17 https://doi.org/10.1007/978-3-319-19168-3_4
- 18 Le, Q. B., Tamene, L., & Vlek, P. L. G. (2012). Multi-pronged assessment of land degradation in
19 West Africa to assess the importance of atmospheric fertilization in masking the processes
20 involved. *Global and Planetary Change*, *92–93*, 71–81.
21 <https://doi.org/10.1016/J.GLOPLACHA.2012.05.003>
- 22 Le, Q., Biradar, C., Thomas, R., Zucca, C., & Bonaiuti, E. (2016). Socio-ecological Context Typology
23 to Support Targeting and Upscaling of Sustainable Land Management Practices in Diverse
24 Global Dryland. *International Congress on Environmental Modelling and Software*. Retrieved
25 from <https://scholarsarchive.byu.edu/iemssconference/2016/Stream-A/45>
- 26 le Roux, B., van der Laan, M., Vahrmeijer, T., Bristow, K. L., & Annandale, J. G. (2017).
27 Establishing and testing a catchment water footprint framework to inform sustainable irrigation
28 water use for an aquifer under stress. *Science of The Total Environment*, *599–600*, 1119–1129.
29 <https://doi.org/10.1016/J.SCITOTENV.2017.04.170>
- 30 Leal, I. R., Da Silva, J. M. C., Tabarelli, M., & Lacher, T. E. (2005). Changing the Course of
31 Biodiversity Conservation in the Caatinga of Northeastern Brazil. *Conservation Biology*, *19*(3),
32 701–706. <https://doi.org/10.1111/j.1523-1739.2005.00703.x>
- 33 Lechmere-Oertel, R. G., Kerley, G. I. H., & Cowling, R. M. (2005). Patterns and implications of
34 transformation in semi-arid succulent thicket, South Africa. *Journal of Arid Environments*, *62*,
35 459–474.
- 36 Lee, H., Honda, Y., Lim, Y.-H., Guo, Y. L., Hashizume, M., & Kim, H. (2014). Effect of Asian dust
37 storms on mortality in three Asian cities. *Atmospheric Environment*, *89*, 309–317.
38 <https://doi.org/10.1016/J.ATMOSENV.2014.02.048>
- 39 Lee, S.-J., & Berbery, E. H. (2012). Land Cover Change Effects on the Climate of the La Plata Basin.
40 *Journal of Hydrometeorology*, *13*(1), 84–102. <https://doi.org/10.1175/JHM-D-11-021.1>
- 41 Leggett, J., Pepper, W. J., Swart, R. J., Edmonds, J., Meira Filho, L. G., Mintzer, I., ... Wasson, J.
42 (1992). Emissions scenarios for the IPCC: an update. *Climate Change 1992: The Supplementary*
43 *Report to the IPCC Scientific Assessment*, 69–95. <https://doi.org/10.1007/s00213-003-1546-3>
- 44 Lei, Y., Hoskins, B., & Slingo, J. (2011). Exploring the interplay between natural decadal variability
45 and anthropogenic climate change in summer rainfall over China. Part I: Observational
46 evidence. *Journal of Climate*, *24*(17), 4584–4599. <https://doi.org/10.1175/2010JCLI3794.1>
- 47 Lemordant, L., Gentine, P., Swann, A. S., Cook, B. I., & Scheff, J. (2018). Critical impact of
48 vegetation physiology on the continental hydrologic cycle in response to increasing CO₂.

- 1 *Proceedings of the National Academy of Sciences of the United States of America*, 115(16),
2 4093–4098. <https://doi.org/10.1073/pnas.1720712115>
- 3 Lenes, J. M., Darrow, B. P., Cattrall, C., Heil, C. A., Callahan, M., Vargo, G. A., ... Walsh, J. J.
4 (2001). Iron fertilization and the Trichodesmium response on the West Florida shelf. *Limnology*
5 *and Oceanography*, 46(6), 1261–1277. <https://doi.org/10.4319/lo.2001.46.6.1261>
- 6 Leroux, L., Bégué, A., Lo Seen, D., Jolivot, A., & Kayitakire, F. (2017). Driving forces of recent
7 vegetation changes in the Sahel: Lessons learned from regional and local level analyses. *Remote*
8 *Sensing of Environment*, 191, 38–54. <https://doi.org/10.1016/j.rse.2017.01.014>
- 9 Lévesque, M., Siegwolf, R., Saurer, M., Eilmann, B., & Rigling, A. (2014). Increased water-use
10 efficiency does not lead to enhanced tree growth under xeric and mesic conditions. *New*
11 *Phytologist*, 203(1), 94–109. <https://doi.org/10.1111/nph.12772>
- 12 Li, A., Wu, J., & Huang, J. (2012). Distinguishing between human-induced and climate-driven
13 vegetation changes: a critical application of RESTREND in inner Mongolia. *Landscape*
14 *Ecology*, 27(7), 969–982. <https://doi.org/10.1007/s10980-012-9751-2>
- 15 Li, D., Wu, S., Liu, L., Zhang, Y., & Li, S. (2018). Vulnerability of the global terrestrial ecosystems
16 to climate change. *Global Change Biology*, 24(9), 4095–4106.
17 <https://doi.org/10.1111/gcb.14327>
- 18 Li, H., Shen, H., Chen, L., Liu, T., Hu, H., Zhao, X., ... Fang, J. (2016). Effects of shrub
19 encroachment on soil organic carbon in global grasslands. *Scientific Reports*, 6(1), 28974.
20 <https://doi.org/10.1038/srep28974>
- 21 Li, Y., Awada, T., Zhou, X., Shang, W., Chen, Y., Zuo, X., ... Feng, J. (2012). Mongolian pine
22 plantations enhance soil physico-chemical properties and carbon and nitrogen capacities in semi-
23 arid degraded sandy land in China. *Applied Soil Ecology*, 56, 1–9.
24 <https://doi.org/10.1016/J.APSSOIL.2012.01.007>
- 25 Li, Y., Kalnay, E., Motesharrei, S., Rivas, J., Kucharski, F., Kirk-Davidoff, D., ... Zeng, N. (2018).
26 Climate model shows large-scale wind and solar farms in the Sahara increase rain and
27 vegetation. *Science (New York, N.Y.)*, 361(6406), 1019–1022.
28 <https://doi.org/10.1126/science.aar5629>
- 29 Li, Z., & Fang, H. (2016). Impacts of climate change on water erosion: A review. *Earth-Science*
30 *Reviews*, 163, 94–117. <https://doi.org/10.1016/J.EARSCIREV.2016.10.004>
- 31 Liao, C., Barrett, C., & Kassam, K.-A. (2015). Does Diversification Improve Livelihoods? Pastoral
32 Households in Xinjiang, China. *Development and Change*, 46(6), 1302–1330.
33 <https://doi.org/10.1111/dech.12201>
- 34 Liao, C., Clark, P. E., DeGloria, S. D., & Barrett, C. B. (2017). Complexity in the spatial utilization of
35 rangelands: Pastoral mobility in the Horn of Africa. *Applied Geography*, 86, 208–219.
36 <https://doi.org/10.1016/J.APGEOG.2017.07.003>
- 37 Liehr, S., Drees, L., & Hummel, D. (2016). Migration as Societal Response to Climate Change and
38 Land Degradation in Mali and Senegal. In *Adaptation to Climate Change and Variability in*
39 *Rural West Africa* (pp. 147–169). Cham: Springer International Publishing.
40 https://doi.org/10.1007/978-3-319-31499-0_9
- 41 Lim, W.-Y., & Seow, A. (2012). Biomass fuels and lung cancer. *Respirology*, 17(1), 20–31.
42 <https://doi.org/10.1111/j.1440-1843.2011.02088.x>
- 43 Lima, M., Christie, D. A., Santoro, M. C., & Latorre, C. (2016). Coupled Socio-Environmental
44 Changes Triggered Indigenous Aymara Depopulation of the Semiarid Andes of Tarapacá-Chile
45 during the Late 19th-20th Centuries. *PloS One*, 11(8), e0160580.
46 <https://doi.org/10.1371/journal.pone.0160580>
- 47 Lin, L., Gettelman, A., Feng, S., & Fu, Q. (2015). Simulated climatology and evolution of aridity in
48 the 21st century. *Journal of Geophysical Research: Atmospheres*, 120(12), 5795–5815.

- 1 <https://doi.org/10.1002/2014JD022912>
- 2 Lin, M., Chen, C., Wang, Q., Cao, Y., Shih, J., Lee, Y., ... Wang, S. (2009). Fuzzy model-based
3 assessment and monitoring of desertification using MODIS satellite imagery. *Engineering*
4 *Computations*, 26(7), 745–760. <https://doi.org/10.1108/02644400910985152>
- 5 Ling, H., Xu, H., Fu, J., Fan, Z., & Xu, X. (2013). Suitable oasis scale in a typical continental river
6 basin in an arid region of China: A case study of the Manas River Basin. *Quaternary*
7 *International*, 286, 116–125. <https://doi.org/10.1016/j.quaint.2012.07.027>
- 8 Liniger, H., & Critchley, W. (2007). *Case studies and analysis of soil and water conservation*
9 *initiatives worldwide*. University of Minnesota.
- 10 Liniger, H., Mektaschi Studer, R., Moll, P., & Zander, U. (2017). *Making sense of research for*
11 *sustainable land management*.
- 12 Linke, A. M., O’Loughlin, J., McCabe, J. T., Tir, J., & Witmer, F. D. W. (2015). Rainfall variability
13 and violence in rural Kenya: Investigating the effects of drought and the role of local institutions
14 with survey data. *Global Environmental Change*, 34, 35–47.
15 <https://doi.org/10.1016/J.GLOENVCHA.2015.04.007>
- 16 Lioubimtseva, E. (2015). A multi-scale assessment of human vulnerability to climate change in the
17 Aral Sea basin. *Environmental Earth Sciences*, 73(2), 719–729. [https://doi.org/10.1007/s12665-](https://doi.org/10.1007/s12665-014-3104-1)
18 [014-3104-1](https://doi.org/10.1007/s12665-014-3104-1)
- 19 Littell, J., McKenzie, D., Wan, H. Y., & Cushman, S. (2018). *Climate change and future wildfire in*
20 *the western USA: an ecological approach to non-stationarity*. *Earth’s Future* (Vol. 6).
21 <https://doi.org/10.1029/2018EF000878>
- 22 Little, P. (2010). Unofficial Cross-Border Trade in Eastern Africa. In A. Sarris, T. Jayne, & J.
23 Morrison (Eds.), *Food Security in Africa: Market and Trade Policy for Staple Foods in Eastern*
24 *and Southern Africa* (pp. 158–181). Cheltenham, UK: Edwin Elgar Publishing.
- 25 Liu, G., Wang, H., Cheng, Y., Zheng, B., & Lu, Z. (2016a). The impact of rural out-migration on
26 arable land use intensity: Evidence from mountain areas in Guangdong, China. *Land Use Policy*,
27 59, 569–579. <https://doi.org/10.1016/J.LANDUSEPOL.2016.10.005>
- 28 Liu, H.-L., Willems, P., Bao, A.-M., Wang, L., & Chen, X. (2016b). Effect of climate change on the
29 vulnerability of a socio-ecological system in an arid area. *Global and Planetary Change*, 137, 1–
30 9. <https://doi.org/10.1016/J.GLOPLACHA.2015.12.014>
- 31 Liu, J., Liu, Q., & Yang, H. (2016c). Assessing water scarcity by simultaneously considering
32 environmental flow requirements, water quantity, and water quality. *Ecological Indicators*, 60,
33 434–441. <https://doi.org/10.1016/J.ECOLIND.2015.07.019>
- 34 Liu, J., Yang, H., Gosling, S. N., Kummu, M., Flörke, M., Pfister, S., ... Oki, T. (2017). Water
35 scarcity assessments in the past, present, and future. *Earth’s Future*, 5(6), 545–559.
36 <https://doi.org/10.1002/2016EF000518>
- 37 Liu, M., & Tian, H. (2010). China’s land cover and land use change from 1700 to 2005: Estimations
38 from high-resolution satellite data and historical archives. *Global Biogeochemical Cycles*, 24(3),
39 n/a-n/a. <https://doi.org/10.1029/2009GB003687>
- 40 Liu, Y., Liu, J., & Zhou, Y. (2017). Spatio-temporal patterns of rural poverty in China and targeted
41 poverty alleviation strategies. *Journal of Rural Studies*, 52, 66–75.
42 <https://doi.org/10.1016/J.JRURSTUD.2017.04.002>
- 43 Liu, Y. Y., Evans, J. P., McCabe, M. F., de Jeu, R. A. M., van Dijk, A. I. J. M., Dolman, A. J., &
44 Saizen, I. (2013). Changing Climate and Overgrazing Are Decimating Mongolian Steppes. *PLoS*
45 *ONE*, 8(2). <https://doi.org/10.1371/journal.pone.0057599>
- 46 Liu, Z., & Wimberly, M. C. (2016). Direct and indirect effects of climate change on projected future
47 fire regimes in the western United States. *Science of the Total Environment*, 542, 65–75.

- 1 <https://doi.org/10.1016/j.scitotenv.2015.10.093>
- 2 Lockyer, Z. B., Coates, P. S., Casazza, M. L., Espinosa, S., & Delehanty, D. J. (2015). Nest-site
3 selection and reproductive success of greater sage-grouse in a fire-affected habitat of
4 northwestern Nevada. *The Journal of Wildlife Management*, 79(5), 785–797.
5 <https://doi.org/10.1002/jwmg.899>
- 6 Lohmann, S., & Lechtenfeld, T. (2015). The Effect of Drought on Health Outcomes and Health
7 Expenditures in Rural Vietnam. *World Development*, 72, 432–448.
8 <https://doi.org/10.1016/J.WORLDDEV.2015.03.003>
- 9 Lokonon, B. O. K., & Mbaye, A. A. (2018). Climate change and adoption of sustainable land
10 management practices in the Niger basin of Benin. *Natural Resources Forum*, 42(1), 42–53.
11 <https://doi.org/10.1111/1477-8947.12142>
- 12 London Government Office for Science & Foresight. (2011). *Migration and global environmental*
13 *change: future challenges and opportunities*. Retrieved from
14 [https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/](https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/287717/11-1116-migration-and-global-environmental-change.pdf)
15 [287717/11-1116-migration-and-global-environmental-change.pdf](https://assets.publishing.service.gov.uk/government/uploads/system/uploads/attachment_data/file/287717/11-1116-migration-and-global-environmental-change.pdf)
- 16 López-Bermúdez, F. (1990). Soil erosion by water on the desertification of a semi-arid Mediterranean
17 fluvial basin: the Segura basin, Spain. *Agriculture, Ecosystems and Environment*, 33(2), 129–
18 145. [https://doi.org/10.1016/0167-8809\(90\)90238-9](https://doi.org/10.1016/0167-8809(90)90238-9)
- 19 Lopez-Garcia, J., Pozza, A., & Sample, T. (2016). Long-term soiling of silicon PV modules in a
20 moderate subtropical climate. *Solar Energy*, 130, 174–183.
21 <https://doi.org/10.1016/J.SOLENER.2016.02.025>
- 22 López-i-Gelats, F., Fraser, E. D. G., Morton, J. F., & Rivera-Ferre, M. G. (2016). What drives the
23 vulnerability of pastoralists to global environmental change? A qualitative meta-analysis. *Global*
24 *Environmental Change*, 39, 258–274. <https://doi.org/10.1016/J.GLOENVCHA.2016.05.011>
- 25 Louhaichi, M., & Tastad, A. (2010). The Syrian Steppe: Past Trends, Current Status, and Future
26 Priorities. *Rangelands*, 32, 2–7. <https://doi.org/10.2111/1551-501X-32.2.2>
- 27 Lu, J., Vecchi, G. A., & Reichler, T. (2007). Expansion of the Hadley cell under global warming.
28 *Geophysical Research Letters*, 34(6), L06805. <https://doi.org/10.1029/2006GL028443>
- 29 Luedeling, E., Gebauer, J., & Buerkert, A. (2009). Climate change effects on winter chill for tree
30 crops with chilling requirements on the Arabian Peninsula. *Climatic Change*, 96(1–2), 219–237.
31 <https://doi.org/10.1007/s10584-009-9581-7>
- 32 Lybbert, T. J., Barrett, C. B., Desta, S., & Layne Coppock, D. (2004). Stochastic wealth dynamics and
33 risk management among a poor population*. *The Economic Journal*, 114(498), 750–777.
34 <https://doi.org/10.1111/j.1468-0297.2004.00242.x>
- 35 Ma, X., Zhao, C., Tao, H., Zhu, J., & Kundzewicz, Z. W. (2018). Projections of actual
36 evapotranspiration under the 1.5 °C and 2.0 °C global warming scenarios in sandy areas in
37 northern China. *Science of The Total Environment*, 645, 1496–1508.
38 <https://doi.org/10.1016/j.scitotenv.2018.07.253>
- 39 Mach, K. J., Planton, S., & Stechow, C. von (Eds.). (2014). Climate Change 2014: Synthesis Report.
40 Contribution of Working Groups I, II and III to the Fifth Assessment Report of the
41 Intergovernmental Panel on Climate Change. In *Annex II: Glossary. Climate Change 2014:*
42 *Synthesis Report. Contribution of Working Groups I, II and III to the Fifth Assessment Report of*
43 *the Intergovernmental Panel on Climate Change [Core Writing Team, R.K. Pachauri and L.A.*
44 *Meyer (eds.)*. (pp. 117–130). Geneva, Switzerland.
- 45 Macheche, C., Molle, N., Ayisi, K., Mashatola, M., Anim, F., & Vanasche, F. (2004). *Smallholder*
46 *Irrigation and Agricultural Development in the Olifants River Basin of Limpopo Province:*
47 *Management Transfer, Productivity, Profitability and Food Security Issues*. Retrieved from
48 file:///C:/Users/hpaulos/Downloads/Smallholder_Irrigation_and_Agricultural_Developmen.pdf

- 1 Madani, K., AghaKouchak, A., & Mirchi, A. (2016). Iran's Socio-economic Drought: Challenges of a
2 Water-Bankrupt Nation. *Iranian Studies*, 49(6), 997–1016.
3 <https://doi.org/10.1080/00210862.2016.1259286>
- 4 Maestre, F. T., & Escudero, A. (2009). Is the patch size distribution of vegetation a suitable indicator
5 of desertification processes? *Ecology*, 90(7), 1729–1735. <https://doi.org/10.1890/08-2096.1>
- 6 Maestre, F. T., Quero, J. L., Gotelli, N. J., Escudero, A., Ochoa, V., Delgado-baquerizo, M., ...
7 Cabrera, O. (2012). Plant Species Richness and Ecosystems Multifunctionality in Global
8 Drylands. *Science*, 335(6065), 2014–2017. <https://doi.org/10.1126/science.1215442>
- 9 Magandana, T. P. (2016). *Effect of Acacia karoo encroachment on grass production in the semi-arid*
10 *savannas of the Eastern Cape, South Africa*. University of Fort Hare.
- 11 Maghrabi, A., Alharbi, B., & Tapper, N. (2011). Impact of the March 2009 dust event in Saudi Arabia
12 on aerosol optical properties, meteorological parameters, sky temperature and emissivity.
13 *Atmospheric Environment*, 45(13), 2164–2173.
14 <https://doi.org/10.1016/J.ATMOSENV.2011.01.071>
- 15 Mahmud, W. E., & Watanabe, K. (2014). Modified Grey Model and its application to groundwater
16 flow analysis with limited hydrogeological data: A case study of the Nubian Sandstone, Kharga
17 Oasis, Egypt. *Environmental Monitoring and Assessment*, 186(2), 1063–1081.
18 <https://doi.org/10.1007/s10661-013-3439-1>
- 19 Mahmood, R., Pielke, R. A., & McAlpine, C. A. (2016). Climate-relevant land use and land cover
20 change policies. *Bulletin of the American Meteorological Society*, 97(2), 195–202.
21 <https://doi.org/10.1175/BAMS-D-14-00221.1>
- 22 Mahowald, N. M. (2007). Anthropocene changes in desert area: Sensitivity to climate model
23 predictions. *Geophysical Research Letters*, 34(18), L18817.
24 <https://doi.org/10.1029/2007GL030472>
- 25 Mahowald, N. M., Muhs, D. R., Levis, S., Rasch, P. J., Yoshioka, M., Zender, C. S., & Luo, C.
26 (2006). Change in atmospheric mineral aerosols in response to climate: Last glacial period,
27 preindustrial, modern, and doubled carbon dioxide climates. *Journal of Geophysical Research:*
28 *Atmospheres*, 111(D10), n/a-n/a. <https://doi.org/10.1029/2005JD006653>
- 29 Mainali, K. P., Warren, D. L., Dhileepan, K., McConnachie, A., Strathie, L., Hassan, G., ...
30 Parmesan, C. (2015). Projecting future expansion of invasive species: comparing and improving
31 methodologies for species distribution modeling. *Global Change Biology*, 21(12), 4464–4480.
32 <https://doi.org/10.1111/gcb.13038>
- 33 Mainguet, M., & Dumay, F. (2011). *Fighting wind erosion. one aspect of the combat against*
34 *desertification*. (CSFD/Agropolis International, Ed.). Montpellier, France.: Les dossiers
35 thématiques du CSFD. Retrieved from [http://www.csf-](http://www.csf-desertification.org/bibliotheque/item/library-fighting-wind-erosion)
36 [desertification.org/bibliotheque/item/library-fighting-wind-erosion](http://www.csf-desertification.org/bibliotheque/item/library-fighting-wind-erosion)
- 37 Majeed, A., & Muhammad, Z. (2019). Salinity: A Major Agricultural Problem—Causes, Impacts on
38 Crop Productivity and Management Strategies. In *Plant Abiotic Stress Tolerance* (pp. 83–99).
39 Cham: Springer International Publishing. https://doi.org/10.1007/978-3-030-06118-0_3
- 40 Makondo, C. C., & Thomas, D. S. G. (2018). Climate change adaptation: Linking indigenous
41 knowledge with western science for effective adaptation. *Environmental Science & Policy*, 88,
42 83–91. <https://doi.org/10.1016/J.ENVSCI.2018.06.014>
- 43 Makonese, T., Ifegbesan, A., & Rampedi, I. (2017). *Household cooking fuel use patterns and*
44 *determinants across southern Africa: Evidence from the demographic and health survey data.*
45 *Energy & Environment* (Vol. 29). <https://doi.org/10.1177/0958305X17739475>
- 46 Malcolm, J. R., Liu, C., Neilson, R. P., Hansen, L., & Hannah, L. (2006). Global warming and
47 extinctions of endemic species from biodiversity hotspots. *Conservation Biology*, 20(2), 538–
48 548. <https://doi.org/10.1111/j.1523-1739.2006.00364.x>

- 1 Maldonado, J., Bennett, T. M. B., Chief, K., Cochran, P., Cozzetto, K., Gough, B., ... Voggeser, G.
2 (2016). Engagement with indigenous peoples and honoring traditional knowledge systems.
3 *Climatic Change*, 135(1), 111–126. <https://doi.org/10.1007/s10584-015-1535-7>
- 4 Maliszewski, P. J., Larson, E. K., & Perrings, C. (2012). Environmental determinants of unscheduled
5 residential outages in the electrical power distribution of Phoenix, Arizona. *Reliability*
6 *Engineering & System Safety*, 99, 161–171. <https://doi.org/10.1016/J.RESS.2011.10.011>
- 7 Maliva, R., & Missimer, T. (2012). Aridity and Drought (pp. 21–39). https://doi.org/10.1007/978-3-642-29104-3_2
- 9 Máñez Costa, M. A., Moors, E. J., & Fraser, E. D. G. (2011). Socioeconomics, policy, or climate
10 change: What is driving vulnerability in southern Portugal? *Ecology and Society*, 16(1).
11 <https://doi.org/10.5751/ES-03703-160128>
- 12 Mani, M., & Pillai, R. (2010). Impact of dust on solar photovoltaic (PV) performance: Research
13 status, challenges and recommendations. *Renewable and Sustainable Energy Reviews*, 14(9),
14 3124–3131. <https://doi.org/10.1016/J.RSER.2010.07.065>
- 15 Manjoro, M., Kakembo, V., & Rowntree, K. M. (2012). Trends in soil erosion and woody shrub
16 encroachment in ngqushwa district, eastern cape province, South Africa. *Environmental*
17 *Management*, 49(3), 570–579. <https://doi.org/10.1007/s00267-012-9810-0>
- 18 Manlosa, A. O., Schultner, J., Dorresteijn, I., & Fischer, J. (2018). Leverage points for improving
19 gender equality and human well-being in a smallholder farming context. *Sustainability Science*,
20 1–13. <https://doi.org/10.1007/s11625-018-0636-4>
- 21 Manson, A. D., Jewitt, D., & Short, A. D. (2007). Effects of season and frequency of burning on soils
22 and landscape functioning in a moist montane grassland. *African Journal of Range and Forage*
23 *Science*, 24(1), 9–18.
- 24 Manzano, M. G., & Návar, J. (2000). Processes of desertification by goats overgrazing in the
25 Tamaulipan thornscrub (matorral) in north-eastern Mexico. *Journal of Arid Environments*,
26 44(1), 1–17. <https://doi.org/10.1006/JARE.1999.0577>
- 27 Mao, J., Ribes, A., Yan, B., Shi, X., Thornton, P. E., Séférian, R., ... Lian, X. (2016). Human-induced
28 greening of the northern extratropical land surface. *Nature Climate Change*, 6(10), 959–963.
29 <https://doi.org/10.1038/nclimate3056>
- 30 Mao, J., Shi, X., Thornton, P., Hoffman, F., Zhu, Z., & Myneni, R. (2013). Global Latitudinal-
31 Asymmetric Vegetation Growth Trends and Their Driving Mechanisms: 1982–2009. *Remote*
32 *Sensing*, 5(3), 1484–1497. <https://doi.org/10.3390/rs5031484>
- 33 Mapfumo, P., Mtambanengwe, F., & Chikowo, R. (2016). Building on indigenous knowledge to
34 strengthen the capacity of smallholder farming communities to adapt to climate change and
35 variability in southern Africa. *Climate and Development*, 8(1), 72–82.
36 <https://doi.org/10.1080/17565529.2014.998604>
- 37 Marengo, J. A., & Bernasconi, M. (2015). Regional differences in aridity/drought conditions over
38 Northeast Brazil: present state and future projections. *Climatic Change*, 129(1–2), 103–115.
39 <https://doi.org/10.1007/s10584-014-1310-1>
- 40 Marlet, S., Bouksila, F., & Bahri, A. (2009). Water and salt balance at irrigation scheme scale: A
41 comprehensive approach for salinity assessment in a Saharan oasis. *Agricultural Water*
42 *Management*, 96(9), 1311–1322. <https://doi.org/10.1016/J.AGWAT.2009.04.016>
- 43 Marques, M., Schwilch, G., Lauterburg, N., Crittenden, S., Tesfai, M., Stolte, J., ... Doko, A. (2016).
44 Multifaceted Impacts of Sustainable Land Management in Drylands: A Review. *Sustainability*,
45 8(2), 177. <https://doi.org/10.3390/su8020177>
- 46 Marshall, M., Funk, C., & Michaelsen, J. (2012). Examining evapotranspiration trends in Africa.
47 *Climate Dynamics*, 38(9–10), 1849–1865. <https://doi.org/10.1007/s00382-012-1299-y>

- 1 Marshall, V. M., Lewis, M. M., & Ostendorf, B. (2012). Buffel grass (*Cenchrus ciliaris*) as an invader
2 and threat to biodiversity in arid environments: A review. *Journal of Arid Environments*, 78, 1–
3 12. <https://doi.org/10.1016/J.JARIDENV.2011.11.005>
- 4 Marticorena, B., Chatenet, B., Rajot, J. L., Traoré, S., Coulibaly, M., Diallo, A., ... Zakou, A. (2010).
5 Temporal variability of mineral dust concentrations over West Africa: analyses of a pluriannual
6 monitoring from the AMMA Sahelian Dust Transect. *Atmospheric Chemistry and Physics*,
7 10(18), 8899–8915. <https://doi.org/10.5194/acp-10-8899-2010>
- 8 Martinez-Mena, M., Lopez, J., Almagro, M., Boix-Fayos, C., & Albaladejo, J. (2008). Effect of water
9 erosion and cultivation on the soil carbon stock in a semiarid area of South-East Spain. *Soil and*
10 *Tillage Research*, 99(1), 119–129. <https://doi.org/10.1016/j.still.2008.01.009>
- 11 Martínez-Palacios, A., Eguiarte, L. E., & Furnier, G. R. (1999). Genetic diversity of the endangered
12 endemic *Agave victoriae-reginae* (Agavaceae) in the Chihuahuan Desert. *American Journal of*
13 *Botany*, 86(8), 1093–1098. <https://doi.org/10.2307/2656971>
- 14 Martínez, G. I., Ovalle, C., Pozo, A. ., Uribe, H., Valderrama, V. N., Prat, C., ... Zagal, E. (2011).
15 Influence of Conservation Tillage and Soil Water Content on Crop Yield in Dryland Compacted
16 Alfisol of Central Chile. *Chilean Journal of Agricultural Research*, 71(December), 615–622.
17 <https://doi.org/10.4067/S0718-58392011000400018>
- 18 Marty, J. (2005). Effects of Cattle Grazing on Diversity in Ephemeral Wetlands. *Conservation*
19 *Biology*, 19(5), 1626–1632. <https://doi.org/10.1111/j.1523-1739.2005.00198.x>
- 20 Marwat, K. B., Hashim, S., & Ali, H. (2010). Weed management: A case study from north-west
21 Pakistan. *Pakistan Journal of Botany*, 42(SPEC. ISSUE), 341–353.
- 22 Masih, I., Maskey, S., Mussá, F. E. F., & Trambauer, P. (2014). A review of droughts on the African
23 continent: A geospatial and long-term perspective. *Hydrology and Earth System Sciences*, 18(9),
24 3635–3649. <https://doi.org/10.5194/hess-18-3635-2014>
- 25 Masoud, A. A., & Koike, K. (2006). Arid land salinization detected by remotely-sensed landcover
26 changes: A case study in the Siwa region, NW Egypt. *Journal of Arid Environments*, 66(1),
27 151–167. <https://doi.org/10.1016/J.JARIDENV.2005.10.011>
- 28 Massey, D. S., Axinn, W. G., & Ghimire, D. J. (2010). Environmental change and out-migration:
29 evidence from Nepal. *Population and Environment*, 32(2–3), 109–136.
30 <https://doi.org/10.1007/s11111-010-0119-8>
- 31 Mavlyanova, N., Kulov, K., & Jooshov, P. (2016). *Suggested measures to reduce irrigation erosion in*
32 *Kyrgyzstan Food Security in the Eurasian Region, 2016. (in Russian)*. Retrieved from
33 <https://ecommons.cornell.edu/handle/1813/55721>
- 34 Mays, L. W. (2013). Groundwater Resources Sustainability: Past, Present, and Future. *Water*
35 *Resources Management*, 27(13), 4409–4424. <https://doi.org/10.1007/s11269-013-0436-7>
- 36 Maystadt, J.-F., & Ecker, O. (2014). Extreme Weather and Civil War: Does Drought Fuel Conflict in
37 Somalia through Livestock Price Shocks? *American Journal of Agricultural Economics*, 96(4),
38 1157–1182. <https://doi.org/10.1093/ajae/aau010>
- 39 Mbow, C. (2017). *The Great Green Wall in the Sahel* (Vol. 1). Oxford University Press.
40 <https://doi.org/10.1093/acrefore/9780190228620.013.559>
- 41 McCaw, W. L. (2013). Managing forest fuels using prescribed fire – A perspective from southern
42 Australia. *Forest Ecology and Management*, 294, 217–224.
43 <https://doi.org/https://doi.org/10.1016/j.foreco.2012.09.012>
- 44 McConnachie, A. J., Strathie, L. W., Mersie, W., Gebrehiwot, L., Zewdie, K., & Abdurehim, A.
45 (2011). Current and potential geographical distribution of the invasive plant *Parthenium*
46 *hysterophorus* (Asteraceae) in eastern and southern Africa. *Weed Research*, 51(1), 71–84.
47 <https://doi.org/10.1111/j.1365-3180.2010.00820.x>

- 1 McCord, P., Dell'Angelo, J., Baldwin, E., & Evans, T. (2017). Polycentric Transformation in Kenyan
2 Water Governance: A Dynamic Analysis of Institutional and Social-Ecological Change. *Policy*
3 *Studies Journal*, 45(4), 633–658. <https://doi.org/10.1111/psj.12168>
- 4 McKenzie, D., & Yang, D. (2015). Evidence on Policies to Increase the Development Impacts of
5 International Migration. *The World Bank Research Observer*, 30(2), 155–192.
6 <https://doi.org/10.1093/wbro/lkv001>
- 7 McKinley, D. C., Miller-Rushing, A. J., Ballard, H. L., Bonney, R., Brown, H., Cook-Patton, S. C., ...
8 Soukup, M. A. (2017). Citizen science can improve conservation science, natural resource
9 management, and environmental protection. *Biological Conservation*, 208, 15–28.
10 <https://doi.org/10.1016/J.BIOCON.2016.05.015>
- 11 McLeman, R. (2013). Developments in modelling of climate change-related migration. *Climatic*
12 *Change*, 117(3), 599–611. <https://doi.org/10.1007/s10584-012-0578-2>
- 13 McLeman, R. (2017). *Migration and Land Degradation: Recent experience and future trends* (Global
14 Land Outlook Working Paper). Retrieved from
15 [https://knowledge.unccd.int/sites/default/files/2018-06/8.](https://knowledge.unccd.int/sites/default/files/2018-06/8.Migration%2Band%2BLand%2BDegradation__R_McLeman.pdf)
16 [Migration%2Band%2BLand%2BDegradation__R_McLeman.pdf](https://knowledge.unccd.int/sites/default/files/2018-06/8.Migration%2Band%2BLand%2BDegradation__R_McLeman.pdf)
- 17 McLeman, R. A. (2011). Settlement abandonment in the context of global environmental change.
18 *Global Environmental Change*, 21, S108–S120.
19 <https://doi.org/10.1016/J.GLOENVCHA.2011.08.004>
- 20 McPeak, J., Little, P., & Demment, M. (2006). Policy Implications and Future Research Directions. In
21 J. McPeak & P. Little (Eds.), *Pastoral Livestock Marketing in Eastern Africa* (pp. 247–256).
22 Bourton Hall: Intermediate Technology Publications.
23 <https://doi.org/10.3362/9781780440323.014>
- 24 MEA. (2005). *Ecosystems and Human Well-being: Desertification Synthesis*. Washington D.C.:
25 World Resources Institute.
- 26 Megnounif, A., & Ghenim, A. N. (2016). Rainfall irregularity and its impact on the sediment yield in
27 Wadi Sebdou watershed, Algeria. *Arabian Journal of Geosciences*, 9(4), 267.
28 <https://doi.org/10.1007/s12517-015-2280-y>
- 29 Meijer, E., Querner, E., & Boesveld, H. (2013). *Impact of farm dams on river flows; A case study in*
30 *the Limpopo River basin, Southern Africa*. Wageningen.
- 31 Mejia, F. A., & Kleissl, J. (2013). Soiling losses for solar photovoltaic systems in California. *Solar*
32 *Energy*, 95, 357–363. <https://doi.org/10.1016/J.SOLENER.2013.06.028>
- 33 Mejia, F., Kleissl, J., & Bosch, J. L. (2014). The Effect of Dust on Solar Photovoltaic Systems.
34 *Energy Procedia*, 49, 2370–2376. <https://doi.org/10.1016/J.EGYPRO.2014.03.251>
- 35 Mekonnen, G. (2017). Threats and Management Options of Parthenium (*Parthenium hysterophorus* L.)
36 in Ethiopia. *Agriculture Research and Technology*, 10(5), 1–7.
37 <https://doi.org/10.19080/ARTOAJ.2017.10.555798>
- 38 Mekuria, W., Yami, M., Haile, M., Gebrehiwot, K., & Birhane, E. (2018). Impact of exclosures on
39 wood biomass production and fuelwood supply in northern Ethiopia. *Journal of Forestry*
40 *Research*, 1–9. <https://doi.org/10.1007/s11676-018-0643-4>
- 41 Melde, S., Laczko, F., & Gemenne, F. (2017). *Making Mobility Work for Adaptation to*
42 *Environmental Changes: Results from the MECLEP Global Research*.
- 43 Méndez, M., & Magaña, V. (2010). Regional Aspects of Prolonged Meteorological Droughts over
44 Mexico and Central America. *Journal of Climate*, 23(5), 1175–1188.
45 <https://doi.org/10.1175/2009JCLI3080.1>
- 46 Meng, X. H., Evans, J. P., & McCabe, M. F. (2014). The influence of inter-annually varying albedo
47 on regional climate and drought. *Climate Dynamics*, 42(3–4), 787–803.

- 1 <https://doi.org/10.1007/s00382-013-1790-0>
- 2 Meng, X. H., Evans, J. P., McCabe, M. F., Meng, X. H., Evans, J. P., & McCabe, M. F. (2014). The
3 Impact of Observed Vegetation Changes on Land–Atmosphere Feedbacks During Drought.
4 *Journal of Hydrometeorology*, 15(2), 759–776. <https://doi.org/10.1175/JHM-D-13-0130.1>
- 5 Merdas, S., Nouar, B., & Lakhdari, F. (2015). The Green Dam in Algeria as a tool to combat
6 desertification. *Planet@Risk*, 3(1), 3–6. Retrieved from [https://planet-](https://planet-risk.org/index.php/pr/article/view/171/319%0A%0A)
7 [risk.org/index.php/pr/article/view/171/319%0A%0A](https://planet-risk.org/index.php/pr/article/view/171/319%0A%0A)
- 8 Mi, J., Li, J., Chen, D., Xie, Y., & Bai, Y. (2015). Predominant control of moisture on soil organic
9 carbon mineralization across a broad range of arid and semiarid ecosystems on the Mongolia
10 plateau. *Landscape Ecology*, 30(9), 1683–1699. <https://doi.org/10.1007/s10980-014-0040-0>
- 11 Miao, L., Moore, J. C., Zeng, F., Lei, J., Ding, J., He, B., & Cui, X. (2015a). Footprint of Research in
12 Desertification Management in China. *Land Degradation & Development*, 26(5), 450–457.
13 <https://doi.org/10.1002/ldr.2399>
- 14 Miao, L., Ye, P., He, B., Chen, L., & Cui, X. (2015b). Future climate impact on the desertification in
15 the dry land asia using AVHRR GIMMS NDVI3g data. *Remote Sensing*, 7(4), 3863–3877.
16 <https://doi.org/10.3390/rs70403863>
- 17 Micklin, P. (2016). The future Aral Sea: hope and despair. *Environmental Earth Sciences*, 75(9), 844.
18 <https://doi.org/10.1007/s12665-016-5614-5>
- 19 Middleton, N. J. (2017). Desert dust hazards: A global review. *Aeolian Research*, 24, 53–63.
20 <https://doi.org/10.1016/J.AEOLIA.2016.12.001>
- 21 Middleton, N., & Kang, U. (2017). Sand and Dust Storms: Impact Mitigation. *Sustainability*, 9(6),
22 1053. <https://doi.org/10.3390/su9061053>
- 23 Millennium Ecosystem Assessment. (2005). *Ecosystems and Human Well-Being: desertification*
24 *synthesis. Ecosystems and human well-being*. [https://doi.org/ISBN: 1-56973-590-5](https://doi.org/ISBN:1-56973-590-5)
- 25 Miller, R. F., Knick, S. T., Pyke, D. A., Meinke, C. W., Hanser, S. E., Wisdom, M. J., & Hild, A. L.
26 (2011). Characteristics of sagebrush habitats and limitations to long-term conservation. Greater
27 sage-grouse: ecology and conservation of a landscape species and its habitats. *Studies in Avian*
28 *Biology*, 38, 145–184.
- 29 Miller, G., Friedel, M., Adam, P., & Chewings, V. (2010). Ecological impacts of buffel grass
30 (*Cenchrus ciliaris* L.) invasion in central Australia - does field evidence support a fire-invasion
31 feedback? *The Rangeland Journal*, 32(4), 353. <https://doi.org/10.1071/RJ09076>
- 32 Miller, R., Chambers, J., Pyke, D., & Pierson, F. (2013). *A review of fire effects on vegetation and*
33 *soils in the Great Basin Region: response and ecological site characteristics*. Retrieved from
34 [https://www.researchgate.net/profile/Richard_Miller22/publication/259077988_A_reviewof_fire](https://www.researchgate.net/profile/Richard_Miller22/publication/259077988_A_reviewof_fire_effects_on_vegetation_and_soils_in_the_Great_Basin_region_Response_and_site_characteristics/links/558cd34308aee43bf6ae445d.pdf)
35 [_effects_on_vegetation_and_soils_in_the_Great_Basin_region_Response_and_site_characteristi](https://www.researchgate.net/profile/Richard_Miller22/publication/259077988_A_reviewof_fire_effects_on_vegetation_and_soils_in_the_Great_Basin_region_Response_and_site_characteristics/links/558cd34308aee43bf6ae445d.pdf)
36 [cs/links/558cd34308aee43bf6ae445d.pdf](https://www.researchgate.net/profile/Richard_Miller22/publication/259077988_A_reviewof_fire_effects_on_vegetation_and_soils_in_the_Great_Basin_region_Response_and_site_characteristics/links/558cd34308aee43bf6ae445d.pdf)
- 37 Miller, R. L., Tegen, I., & Perlwitz, J. (2004). Surface radiative forcing by soil dust aerosols and the
38 hydrologic cycle. *Journal of Geophysical Research: Atmospheres*, 109(D4), n/a-n/a.
39 <https://doi.org/10.1029/2003JD004085>
- 40 Milly, P. C. D., & Dunne, K. A. (2016). Potential evapotranspiration and continental drying. *Nature*
41 *Climate Change*, 6(10), 946–949. <https://doi.org/10.1038/nclimate3046>
- 42 Milton, S. J. (2003). “Emerging ecosystems” - a washing-stone for ecologists, economists and
43 sociologists? *South African Journal of Science*, 99, 404–406.
- 44 Ministry of Agriculture and Water Resources of Tunisia, & GIZ. (2007). *Stratégie nationale*
45 *d’adaptation de l’agriculture tunisienne et des écosystèmes aux changements climatiques*.
- 46 Ministry of Energy Industry and Mineral Resources. (2016). Third National Communication of the
47 Kingdom of Saudi Arabia, 173–174.

- 1 Mirzabaev, A., Ahmed, M., Werner, J., Pender, J., & Louhaichi, M. (2016). Rangelands of Central
2 Asia: challenges and opportunities. *Journal of Arid Land*, 8(1), 93–108.
3 <https://doi.org/10.1007/s40333-015-0057-5>
- 4 Mirzabaev, A., Goedecke, J., Dubovyk, O., Djanibekov, U., Le, Q. B., & Aw-Hassan, A. (2016).
5 Economics of Land Degradation in Central Asia. In *Economics of Land Degradation and*
6 *Improvement – A Global Assessment for Sustainable Development* (pp. 261–290). Cham:
7 Springer International Publishing. https://doi.org/10.1007/978-3-319-19168-3_10
- 8 Missirian, A., & Schlenker, W. (2017). Asylum applications respond to temperature fluctuations.
9 *Science (New York, N.Y.)*, 358(6370), 1610–1614. <https://doi.org/10.1126/science.aao0432>
- 10 Mo, K. C., & Lettenmaier, D. P. (2015). Heat wave flash droughts in decline. *Geophysical Research*
11 *Letters*, 42(8), 2823–2829. <https://doi.org/10.1002/2015GL064018>
- 12 Molesworth, A. M., Cuevas, L. E., Connor, S. J., Morse, A. P., & Thomson, M. C. (2003).
13 Environmental risk and meningitis epidemics in Africa. *Emerging Infectious Diseases*, 9(10),
14 1287–93. <https://doi.org/10.3201/eid0910.030182>
- 15 Molua, E. L., & Lambi, C. M. (2007). *The Economic Impact Of Climate Change On Agriculture In*
16 *Cameroon, Volume 1 of 1*. The World Bank. <https://doi.org/10.1596/1813-9450-4364>
- 17 Morales, C., Brzovic, F., Dascal, G., Aranibar, Z., Mora, L., Morera, R., Estupiñan, R., Candia, D.,
18 Agar, S., López-Cordovezm, L. and Parada, S. (2011). Measuring the economic value of land
19 degradation/desertification considering the effects of climate change. A study for Latin America
20 and the Caribbean. In *Communication au Séminaire « Politiques, programmes et projets de lutte*
21 *contre la désertification, quelles évaluations ? »*. Montpellier, France.
- 22 Moreno-Calles, A. I., Casas, A., García-Frapolli, E., & Torres-García, I. (2012). Traditional
23 agroforestry systems of multi-crop “milpa” and “chichipera” cactus forest in the arid Tehuacán
24 Valley, Mexico: their management and role in people’s subsistence. *Agroforestry Systems*,
25 84(2), 207–226. <https://doi.org/10.1007/s10457-011-9460-x>
- 26 Moridnejad, A., Karimi, N., & Ariya, P. A. (2015). Newly desertified regions in Iraq and its
27 surrounding areas: Significant novel sources of global dust particles. *Journal of Arid*
28 *Environments*, 116, 1–10. <https://doi.org/10.1016/J.JARIDENV.2015.01.008>
- 29 Moritz, M. A., Parisien, M.-A., Batllori, E., Krawchuk, M. A., Van Dorn, J., Ganz, D. J., & Hayhoe,
30 K. (2012). Climate change and disruptions to global fire activity. *Ecosphere*, 3(6), art49.
31 <https://doi.org/10.1890/ES11-00345.1>
- 32 Mortimore, M. (2005). Dryland development: Success stories from West Africa. *Environment*, 47(1),
33 8–21. <https://doi.org/10.3200/ENVT.47.1.8-21>
- 34 Mortimore, M. (2016). Changing Paradigms for People-Centred Development in the Sahel (pp. 65–
35 98). Springer, Berlin, Heidelberg. https://doi.org/10.1007/978-3-642-16014-1_3
- 36 Morton, J. (2010). Why Should Governmentality Matter for the Study of Pastoral Development?
37 *Nomadic Peoples*, 14(1), 6–30. <https://doi.org/10.3167/np.2010.140102>
- 38 Morton, J. (2017). Climate Change and African Agriculture. In *Making Climate Compatible*
39 *Development Happen* (pp. 87–113). Routledge. <https://doi.org/10.4324/9781315621579-5>
- 40 Mosase, E., & Ahiablame, L. (2018). Rainfall and Temperature in the Limpopo River Basin, Southern
41 Africa: Means, Variations, and Trends from 1979 to 2013. *Water*, 10(4), 364.
42 <https://doi.org/10.3390/w10040364>
- 43 Mosley, L. (2014). Drought impacts on the water quality of freshwater systems; review and
44 integration. *Earth-Science Reviews*, 140.
- 45 Mostephaoui, M., Merdas, S., Sakaa, B., Hanafi, M., & Benazzouz, M. (2013). Cartographie des
46 risques d'érosion hydrique par l'application de l'équation universelle de pertes en sol à l'aide
47 d'un système d'information géographique dans le bassin versant d'El hamel (Boussaâda)

- 1 Algérie [Mapping of water erosion by the application of. *Journal Algérien Des Régions Arides,*
2 *Nspécial*, 131–147. Retrieved from
3 http://www.academia.edu/download/46407430/Cartographie_des_risques_drosion_hydriqu2016
4 [0611-22892-15nohuy.pdf](https://doi.org/10.1175/BAMS-86-1-39)
- 5 Mote, P. W., Hamlet, A. F., Clark, M. P., & Lettenmaier, D. P. (2005). Declining mountain snowpack
6 in western north America. *Bulletin of the American Meteorological Society*, 86(1), 39–49.
7 <https://doi.org/10.1175/BAMS-86-1-39>
- 8 Moussa, B., Nkonya, E., Meyer, S., Kato, E., Johnson, T., & Hawkins, J. (2016). Economics of Land
9 Degradation and Improvement in Niger. In *Economics of Land Degradation and Improvement –*
10 *A Global Assessment for Sustainable Development* (pp. 499–539). Cham: Springer International
11 Publishing. https://doi.org/10.1007/978-3-319-19168-3_17
- 12 Mudd, G. M. (2000). Mound springs of the Great Artesian Basin in South Australia: A case study
13 from Olympic Dam. *Environmental Geology*, 39(5), 463–476.
14 <https://doi.org/10.1007/s002540050452>
- 15 Mueller, V., Gray, C., & Kosec, K. (2014). Heat stress increases long-term human migration in rural
16 Pakistan. *Nature Climate Change*, 4(3), 182–185. <https://doi.org/10.1038/nclimate2103>
- 17 Mulinge, W., Gicheru, P., Murithi, F., Maingi, P., Kihui, E., Kirui, O. K., & Mirzabaev, A. (2016).
18 Economics of Land Degradation and Improvement in Kenya. In *Economics of Land*
19 *Degradation and Improvement – A Global Assessment for Sustainable Development* (pp. 471–
20 498). Cham: Springer International Publishing. https://doi.org/10.1007/978-3-319-19168-3_16
- 21 Müller, B., Johnson, L., & Kreuer, D. (2017). Maladaptive outcomes of climate insurance in
22 agriculture. *Global Environmental Change*, 46, 23–33.
23 <https://doi.org/10.1016/J.GLOENVCHA.2017.06.010>
- 24 Muriithi, B. W., & Matz, J. A. (2015). Welfare effects of vegetable commercialization: Evidence from
25 smallholder producers in Kenya. *Food Policy*, 50, 80–91.
26 <https://doi.org/10.1016/J.FOODPOL.2014.11.001>
- 27 Muyibul, Z., Jianxin, X., Muhtar, P., Qingdong, S., & Run, Z. (2018). Spatiotemporal changes of land
28 use/cover from 1995 to 2015 in an oasis in the middle reaches of the Keriya River, southern
29 Tarim Basin, Northwest China. *CATENA*, 171, 416–425.
30 <https://doi.org/10.1016/J.CATENA.2018.07.038>
- 31 Myhre, G., D. Shindell, F.-M. Breon, W. Collins, J. Fuglestedt, J. Huang, D. Koch, J.-F. Lamarque,
32 D. Lee, B. Mendoza, T. Nakajima, A. Robock, G. Stephens, T. T. and H. Z. (2013).
33 Anthropogenic and Natural Radiative Forcing. In S. K. A. Stocker, T.F., D. Qin, G.-K. Plattner,
34 M. Tignor & V. B. and P. M. M. J. Boschung, A. Nauels, Y. Xia (Eds.), *Climate Change 2013:*
35 *The Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of*
36 *the Intergovernmental Panel on Climate Change*. United Kingdom and New York, NY, USA.:
37 Cambridge University Press, Cambridge. Retrieved from
38 https://www.ipcc.ch/site/assets/uploads/2018/02/WG1AR5_Chapter08_FINAL.pdf
- 39 Mythili, G., & Goedecke, J. (2016). Economics of Land Degradation in India. In *Economics of Land*
40 *Degradation and Improvement – A Global Assessment for Sustainable Development* (pp. 431–
41 469). Cham: Springer International Publishing. https://doi.org/10.1007/978-3-319-19168-3_15
- 42 Nair, P. K. R., Saha, S. K., Nair, V. D., & Haile, S. G. (2011). Potential for greenhouse gas emissions
43 from soil carbon stock following biofuel cultivation on degraded lands. *Land Degradation &*
44 *Development*, 22(4), 395–409. <https://doi.org/10.1002/ldr.1016>
- 45 Nakasone, E., Torero, M., & Minten, B. (2014). The Power of Information: The ICT Revolution in
46 Agricultural Development. *Annual Review of Resource Economics*, 6(1), 533–550.
47 <https://doi.org/10.1146/annurev-resource-100913-012714>
- 48 Namdari, S., Karimi, N., Sorooshian, A., Mohammadi, G., & Sehatkashani, S. (2018). Impacts of
49 climate and synoptic fluctuations on dust storm activity over the Middle East. *Atmospheric*

- 1 *Environment*, 173, 265–276. <https://doi.org/10.1016/J.ATMOSENV.2017.11.016>
- 2 Narayanamoorthy, A. (2010). *Can drip method of irrigation be used to achieve the macro objectives*
3 *of conservation agriculture? Indian Journal of Agricultural Economics* (Vol. 65).
- 4 Nardone, A., Ronchi, B., Lacetera, N., Ranieri, M. S., & Bernabucci, U. (2010). Effects of climate
5 changes on animal production and sustainability of livestock systems. *Livestock Science*, 130(1–
6 3), 57–69. <https://doi.org/10.1016/j.livsci.2010.02.011>
- 7 Nastos, P. T., Politi, N., & Kapsomenakis, J. (2013). Spatial and temporal variability of the Aridity
8 Index in Greece. *Atmospheric Research*, 119, 140–152.
9 <https://doi.org/10.1016/j.atmosres.2011.06.017>
- 10 Naumann, G., Alfieri, L., Wyser, K., Mentaschi, L., Betts, R. A., Carrao, H., ... Feyen, L. (2018).
11 Global Changes in Drought Conditions Under Different Levels of Warming. *Geophysical*
12 *Research Letters*, 45(7), 3285–3296. <https://doi.org/10.1002/2017GL076521>
- 13 Nawrotzki, R. J., & Bakhtsiyarava, M. (2017). International Climate Migration: Evidence for the
14 Climate Inhibitor Mechanism and the Agricultural Pathway. *Population, Space and Place*,
15 23(4), e2033. <https://doi.org/10.1002/psp.2033>
- 16 Nawrotzki, R. J., & DeWaard, J. (2016). Climate shocks and the timing of migration from Mexico.
17 *Population and Environment*, 38(1), 72–100. <https://doi.org/10.1007/s11111-016-0255-x>
- 18 Nawrotzki, R. J., Riosmena, F., Hunter, L. M., & Runfola, D. M. (2015). Undocumented migration in
19 response to climate change. *International Journal of Population Studies*, 1(1), 60–74.
20 <https://doi.org/10.18063/IJPS.2015.01.004>
- 21 Nawrotzki, R. J., Runfola, D. M., Hunter, L. M., & Riosmena, F. (2016). Domestic and International
22 Climate Migration from Rural Mexico. *Human Ecology*, 44(6), 687–699.
23 <https://doi.org/10.1007/s10745-016-9859-0>
- 24 Nearing, M. A., Unkrich, C. L., Goodrich, D. C., Nichols, M. H., & Keefer, T. O. (2015). Temporal
25 and elevation trends in rainfall erosivity on a 149 km² watershed in a semi-arid region of the
26 American Southwest. *International Soil and Water Conservation Research*, 3(2), 77–85.
27 <https://doi.org/10.1016/j.iswcr.2015.06.008>
- 28 Negash, D., Abegaz, A., Smith, J. U., Araya, H., & Gelana, B. (2017). Household energy and
29 recycling of nutrients and carbon to the soil in integrated crop-livestock farming systems: a case
30 study in Kumbursa village, Central Highlands of Ethiopia. *GCB Bioenergy*, 9(10), 1588–1601.
31 <https://doi.org/10.1111/gcbb.12459>
- 32 Nelson, G. C., Rosegrant, M., Palazzo, A., Gray, I., Ingersoll, C., Robertson, R., ... You, L. (2010).
33 *Food security, farming, and climate change to 2050: scenarios, results, policy options*.
34 Washington, D.C.: International Food Policy Research Institute.
- 35 Nelson, V., Forsythe, L., & Morton, J. (2015). *Empowering dryland women: capturing opportunities*
36 *in land rights, governance and resilience. A synthesis of thematic papers from the series*
37 *‘Women’s empowerment in the drylands’*.
- 38 Nerantzaki, S. D., Giannakis, G. V., Efstathiou, D., Nikolaidis, N. P., Sibetheros, I. A., Karatzas, G.
39 P., & Zacharias, I. (2015). Modeling suspended sediment transport and assessing the impacts of
40 climate change in a karstic Mediterranean watershed. *Science of the Total Environment*, 538,
41 288–297. <https://doi.org/10.1016/j.scitotenv.2015.07.092>
- 42 Neuer, S., Torres-Padrón, M. E., Gelado-Caballero, M. D., Rueda, M. J., Hernández-Brito, J.,
43 Davenport, R., & Wefer, G. (2004). Dust deposition pulses to the eastern subtropical North
44 Atlantic gyre: Does ocean’s biogeochemistry respond? *Global Biogeochemical Cycles*, 18(4),
45 n/a-n/a. <https://doi.org/10.1029/2004GB002228>
- 46 Neufeldt, H., Langford, K., Fuller, J., Iiyama, M., & Dobie, P. (2015). *From transition fuel to viable*
47 *energy source: improving sustainability in the sub-Saharan charcoal sector* (ICRAF Working
48 Paper No. 196). Nairobi. <https://doi.org/http://dx.doi.org/10.5716/WP15011.PDF>

- 1 Newbold, T., Hudson, L. N., Hill, S. L. L., Contu, S., Lysenko, I., Senior, R. A., ... Purvis, A. (2015).
2 Global effects of land use on local terrestrial biodiversity. *Nature*, 520(7545), 45–50.
3 <https://doi.org/10.1038/nature14324>
- 4 Newton, A. C., Echeverría, C., Cantarello, E., & Bolados, G. (2011). Projecting impacts of human
5 disturbances to inform conservation planning and management in a dryland forest landscape.
6 *Biological Conservation*, 144(7), 1949–1960. <https://doi.org/10.1016/J.BIOCON.2011.03.026>
- 7 Ngigi, S. N. (2003). What is the limit of up-scaling rainwater harvesting in a river basin? *Physics and*
8 *Chemistry of the Earth, Parts A/B/C*, 28(20–27), 943–956.
9 <https://doi.org/10.1016/J.PCE.2003.08.015>
- 10 NIAB. (1997). *Economic utilization of salt-affected soils In Twenty Five Years (1992-1997) of NIAB*
11 *(Silver Jubilee Publication)*. Nuclear Institute for Agriculture and Biology. Faisalabad, Pakistan:
12 Nuclear Institute for Agriculture and Biology, Faisalabad, Pakistan.
- 13 Niaz, S., Ali, M. A., Ali, S., & Awan, S. (2009). *Comparative Water Use Efficiency of Drip and*
14 *Furrow Irrigation Systems for Off-Season Vegetables under Plastic Tunnel in Rainfed Areas.*
15 *Life Sciences International Journal* (Vol. 2).
- 16 Niedermeyer, E. M., Prange, M., Mulitza, S., Mollenhauer, G., Schefuß, E., & Schulz, M. (2009).
17 Extratropical forcing of Sahel aridity during Heinrich stadials. *Geophysical Research Letters*,
18 36(20), L20707. <https://doi.org/10.1029/2009GL039687>
- 19 Nielsen, D. L., & Brock, M. A. (2009). Modified water regime and salinity as a consequence of
20 climate change: Prospects for wetlands of Southern Australia. *Climatic Change*, 95(3–4), 523–
21 533. <https://doi.org/10.1007/s10584-009-9564-8>
- 22 Nkonya, E., & Anderson, W. (2015). Exploiting provisions of land economic productivity without
23 degrading its natural capital. *Journal of Arid Environments*, 112, 33–43.
24 <https://doi.org/10.1016/J.JARIDENV.2014.05.012>
- 25 Nkonya, E., Anderson, W., Kato, E., Koo, J., Mirzabaev, A., von Braun, J., & Meyer, S. (2016).
26 Global Cost of Land Degradation. In *Economics of Land Degradation and Improvement – A*
27 *Global Assessment for Sustainable Development* (pp. 117–165). Cham: Springer International
28 Publishing. https://doi.org/10.1007/978-3-319-19168-3_6
- 29 Nkonya, E., Johnson, T., Kwon, H. Y., & Kato, E. (2016). Economics of Land Degradation in Sub-
30 Saharan Africa. In von B. J. Nkonya E., Mirzabaev A. (Ed.), *Economics of Land Degradation*
31 *and Improvement – A Global Assessment for Sustainable Development* (pp. 215–259). Cham:
32 Springer International Publishing. https://doi.org/10.1007/978-3-319-19168-3_9
- 33 Nkonya, E., Place, F., Kato, E., & Mwanjololo, M. (2015). Climate Risk Management Through
34 Sustainable Land Management in Sub-Saharan Africa. In *Sustainable Intensification to Advance*
35 *Food Security and Enhance Climate Resilience in Africa* (pp. 75–111). Cham: Springer
36 International Publishing. https://doi.org/10.1007/978-3-319-09360-4_5
- 37 Noojipady, P., Prince, S. D., & Rishmawi, K. (2015). Reductions in productivity due to land
38 degradation in the drylands of the southwestern united states. *Ecosystem Health and*
39 *Sustainability*, 1(8), 1–15. <https://doi.org/10.1890/EHS15-0020.1>
- 40 Notaro, M., Alkolibi, F., Fadda, E., & Bakhrjy, F. (2013). Trajectory analysis of Saudi Arabian dust
41 storms. *Journal of Geophysical Research: Atmospheres*, 118(12), 6028–6043.
42 <https://doi.org/10.1002/jgrd.50346>
- 43 Nurbekov, A., Akramkhanov, A., Kassam, A., Sydyk, D., Ziyadaullaev, Z., & Lamers, J. P. A. (2016).
44 Conservation Agriculture for combating land degradation in Central Asia: A synthesis. *AIMS*
45 *Agriculture and Food*, 1(2), 144–156. <https://doi.org/10.3934/agrfood.2016.2.144>
- 46 Nyagumbo, I., Nyamadzawo, G., & Madembo, C. (2019). Effects of three in-field water harvesting
47 technologies on soil water content and maize yields in a semi-arid region of Zimbabwe.
48 *Agricultural Water Management*, 216, 206–213. <https://doi.org/10.1016/J.AGWAT.2019.02.023>

- 1 Nyamadzawo, G., Wuta, M., Nyamangara, J., & Gumbo, D. (2013). Opportunities for optimization of
2 in-field water harvesting to cope with changing climate in semi-arid smallholder farming areas
3 of Zimbabwe. *SpringerPlus*, 2(1), 1–9. <https://doi.org/10.1186/2193-1801-2-100>
- 4 Nyanga, A., Kessler, A., & Tenge, A. (2016). Key socio-economic factors influencing sustainable
5 land management investments in the West Usambara Highlands, Tanzania. *Land Use Policy*, 51,
6 260–266. <https://doi.org/10.1016/J.LANDUSEPOL.2015.11.020>
- 7 Nyangena, W. (2008). Social determinants of soil and water conservation in rural Kenya.
8 *Environment, Development and Sustainability*, 10(6), 745–767. [https://doi.org/10.1007/s10668-](https://doi.org/10.1007/s10668-007-9083-6)
9 [007-9083-6](https://doi.org/10.1007/s10668-007-9083-6)
- 10 Nyantakyi-Frimpong, H., & Bezner-Kerr, R. (2015). The relative importance of climate change in the
11 context of multiple stressors in semi-arid Ghana. *Global Environmental Change*, 32, 40–56.
12 <https://doi.org/10.1016/J.GLOENVCHA.2015.03.003>
- 13 Nyman, P., Sheridan, G., & Lane, P. N. J. (2010). Synergistic effects of water repellency and
14 macropore flow on the hydraulic conductivity of a burned forest soil, south-east Australia.
15 *Hydrological Processes*, 24(20), 2871–2887. <https://doi.org/10.1002/hyp.7701>
- 16 Nyong, A., Adesina, F., & Osman Elasha, B. (2007). The value of indigenous knowledge in climate
17 change mitigation and adaptation strategies in the African Sahel. *Mitigation and Adaptation*
18 *Strategies for Global Change*, 12(5), 787–797. <https://doi.org/10.1007/s11027-007-9099-0>
- 19 O'Brien, K., Leichenko, R., Kelkar, U., Venema, H., Aandahl, G., Tompkins, H., ... West, J. (2004).
20 Mapping vulnerability to multiple stressors: climate change and globalization in India. *Global*
21 *Environmental Change*, 14(4), 303–313. <https://doi.org/10.1016/J.GLOENVCHA.2004.01.001>
- 22 O'Connor, T. G., Martindale, G., Morris, C. D., Short, A., Witkowski, E. T. F., & Scott-shaw, R.
23 (2011). Influence of Grazing Management on Plant Diversity of Highland Sourveld Influence of
24 Grazing Management on Plant Diversity of Highland Sourveld Grassland ,. *Rangeland Ecology*
25 *& Management*, 64(2), 196–207. <https://doi.org/10.2111/REM-D-10-00062.1>
- 26 O'Connor, T. G., Puttick, J. R., & Hoffman, M. T. (2014). Bush encroachment in southern Africa:
27 changes and causes. *African Journal of Range & Forage Science*, 31(2), 67–88.
28 <https://doi.org/10.2989/10220119.2014.939996>
- 29 Observatoire du Sahara et du Sahel. (2013). La Surveillance environnementale dans le circum-
30 Sahara : Synthèse régionale Ecologie (Algérie , Burkina Faso - Kenya - Mali Niger - Sénégal -
31 Tunisie) 2012 | Observatoire du Sahara et du Sahel. Retrieved May 25, 2018, from
32 [http://www.oss-online.org/fr/la-surveillance-environnementale-dans-le-circum-sahara-synthèse-](http://www.oss-online.org/fr/la-surveillance-environnementale-dans-le-circum-sahara-synthèse-régionale-ecologie-algérie-burkina)
33 [régionale-ecologie-algérie-burkina](http://www.oss-online.org/fr/la-surveillance-environnementale-dans-le-circum-sahara-synthèse-régionale-ecologie-algérie-burkina)
- 34 Odhiambo, G. O. (2017). Water scarcity in the Arabian Peninsula and socio-economic implications.
35 *Applied Water Science*, 7(5), 2479–2492. <https://doi.org/10.1007/s13201-016-0440-1>
- 36 Ogutu, S. O., & Qaim, M. (2019). Commercialization of the small farm sector and multidimensional
37 poverty. *World Development*, 114, 281–293.
38 <https://doi.org/10.1016/J.WORLDDEV.2018.10.012>
- 39 Okhovatian-Ardakani, A. R., Mehrabian, M., Dehghani, F., & Akbarzadeh, A. (2010). Salt
40 tolerance evaluation and relative comparison in cuttings of different pomegranate cultivars.
41 *Plant Soil Environ.*, 2010(4), 176–185. Retrieved from
42 https://www.agriculturejournals.cz/publicFiles/158_2009-PSE.pdf
- 43 Okin, G. S., Baker, A. R., Tegen, I., Mahowald, N. M., Dentener, F. J., Duce, R. A., ... Zhu, T.
44 (2011). Impacts of atmospheric nutrient deposition on marine productivity: Roles of nitrogen,
45 phosphorus, and iron. *Global Biogeochemical Cycles*, 25(2), n/a-n/a.
46 <https://doi.org/10.1029/2010GB003858>
- 47 Olayide, O. E., Tetteh, I. K., & Popoola, L. (2016). Differential impacts of rainfall and irrigation on
48 agricultural production in Nigeria: Any lessons for climate-smart agriculture? *Agricultural*

- 1 *Water Management*, 178, 30–36. <https://doi.org/10.1016/J.AGWAT.2016.08.034>
- 2 Olsson, L. (2017). Climate migration and conflicts. In *Climate Change, Migration and Human Rights*
3 (pp. 116–128). Routledge. <https://doi.org/10.4324/9781315622217-8>
- 4 Olsson, L., Opondo, M., Tschakert, P., Agrawal, A., Eriksen, S. H., Ma, S., ... Zakieldean, S. A.
5 (2014). Livelihoods and poverty. In T. E. B. Field, C.B., V.R. Barros, D.J. Dokken, K.J. Mach,
6 M.D. Mastrandrea, S. M. M. Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova, B. Girma, E.S.
7 Kissel, A.N. Levy, & and L. L. W. P.R. Mastrandrea (Eds.), *Climate Change 2014: Impacts,*
8 *Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working*
9 *Group II to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change.*
10 Cambridge: Cambridge University Press.
- 11 Omuto, C. T., Vargas, R. R., Alim, M. S., & Paron, P. (2010). Mixed-effects modelling of time series
12 NDVI-rainfall relationship for detecting human-induced loss of vegetation cover in drylands.
13 *Journal of Arid Environments*, 74(11), 1552–1563.
14 <https://doi.org/10.1016/J.JARIDENV.2010.04.001>
- 15 Oroda, A. S. (2001). Application Of Remote Sensing To Early Warning For Food Security And
16 Environmental Monitoring In The Horn Of Africa, (Part 6/W6). Retrieved from
17 <https://www.isprs.org/proceedings/xxxiv/6-W6/papers/oroda.pdf>
- 18 Orr, B. J., Cowie, A. L., Castillo, V. M., Chasek, P., Crossman, N. D., Erlewein, A., ... Welton, S.
19 (2017). *Scientific Conceptual Framework for Land Degradation Neutrality. A Report of the*
20 *Science-Policy Interface.* Bonn, Germany. Retrieved from
21 https://www.unccd.int/sites/default/files/documents/2017-08/LDN_CF_report_web-english.pdf
- 22 Osbahr, H., Twyman, C., Neil Adger, W., & Thomas, D. S. G. (2008). Effective livelihood adaptation
23 to climate change disturbance: Scale dimensions of practice in Mozambique. *Geoforum*, 39(6),
24 1951–1964. <https://doi.org/10.1016/J.GEOFORUM.2008.07.010>
- 25 Osgood, D., Powell, B., Diro, R., Farah, C., Enenkel, M., Brown, M., ... McCarty, J. L. (2018).
26 Farmer Perception, Recollection, and Remote Sensing in Weather Index Insurance: An Ethiopia
27 Case Study. *Remote Sensing*, 10(12), 1887. <https://doi.org/10.3390/rs10121887>
- 28 OSS. (2012). *Synthèse Régionale Ecologie(Algérie - Burkina Faso - Kenya - Mali Niger - Sénégal -*
29 *Tunisie) OSS.* Tunis.
- 30 Ostrom, E. (2009). A general framework for analyzing sustainability of social-ecological systems.
31 *Science (New York, N.Y.)*, 325(5939), 419–22. <https://doi.org/10.1126/science.1172133>
- 32 Ostrom, E., & Nagendra, H. (2006). Insights on linking forests, trees, and people from the air, on the
33 ground, and in the laboratory. *Proceedings of the National Academy of Sciences of the United*
34 *States of America*, 103(51), 19224–31. <https://doi.org/10.1073/pnas.0607962103>
- 35 Otto-Bliesner, B. L., Russell, J. M., Clark, P. U., Liu, Z., Overpeck, J. T., Konecky, B., ... Lu, Z.
36 (2014). Coherent changes of southeastern equatorial and northern African rainfall during the last
37 deglaciation. *Science (New York, N.Y.)*, 346(6214), 1223–7.
38 <https://doi.org/10.1126/science.1259531>
- 39 Otuoma, J., Kinyamario, J., Ekaya, W., Kshatriya, M., & Nyabenge, M. (2009). Effects of human-
40 livestock-wildlife interactions on habitat in an eastern Kenya rangeland. *African Journal of*
41 *Ecology*, 47(4), 567–573. <https://doi.org/10.1111/j.1365-2028.2008.01009.x>
- 42 Owain, E. L., & Maslin, M. A. (2018). Assessing the relative contribution of economic, political and
43 environmental factors on past conflict and the displacement of people in East Africa. *Palgrave*
44 *Communications*, 4(1), 47. <https://doi.org/10.1057/s41599-018-0096-6>
- 45 Oweis, T., & Hachum, A. (2006). Water harvesting and supplemental irrigation for improved water
46 productivity of dry farming systems in West Asia and North Africa. *Agricultural Water*
47 *Management*, 80(1–3), 57–73. <https://doi.org/10.1016/J.AGWAT.2005.07.004>
- 48 Oweis, T. Y. (2017). Rainwater harvesting for restoring degraded dry agro-pastoral ecosystems: a

- 1 conceptual review of opportunities and constraints in a changing climate. *Environmental*
2 *Reviews*, 25(2), 135–149. <https://doi.org/10.1139/er-2016-0069>
- 3 Painter, T. H., Flanner, M. G., Kaser, G., Marzeion, B., VanCuren, R. A., & Abdalati, W. (2013). End
4 of the Little Ice Age in the Alps forced by industrial black carbon. *Proceedings of the National*
5 *Academy of Sciences*, 110(38), 15216–15221. <https://doi.org/10.1073/PNAS.1302570110>
- 6 Painter, T. H., Skiles, S. M., Deems, J. S., Brandt, W. T., & Dozier, J. (2018). Variation in Rising
7 Limb of Colorado River Snowmelt Runoff Hydrograph Controlled by Dust Radiative Forcing in
8 Snow. *Geophysical Research Letters*, 45(2), 797–808. <https://doi.org/10.1002/2017GL075826>
- 9 Paltasingh, K. R. (2018). Land tenure security and adoption of modern rice technology in Odisha,
10 Eastern India: Revisiting Besley’s hypothesis. *Land Use Policy*, 78, 236–244.
11 <https://doi.org/10.1016/J.LANDUSEPOL.2018.06.031>
- 12 Panda, R. M., Behera, M. D., & Roy, P. S. (2018). Assessing distributions of two invasive species of
13 contrasting habits in future climate. *Journal of Environmental Management*, 213, 478–488.
14 <https://doi.org/10.1016/J.JENVMAN.2017.12.053>
- 15 Pankova, E. I. (2016). Salinization of Irrigated Soils in the Middle-Asian Region: Old and New
16 Issues. *Arid Ecosystems*, 6(4), 241–248. <https://doi.org/10.1134/S2079096116040077>
- 17 Panthou, G., Vischel, T., & Lebel, T. (2014). Recent trends in the regime of extreme rainfall in the
18 Central Sahel. *International Journal of Climatology*, 34, 3998–4006.
19 <https://doi.org/10.1002/joc.3984>
- 20 Papanastasis, V. P., Bautista, S., Chouvardas, D., Mantzanas, K., Papadimitriou, M., Mayor, A. G., ...
21 Vallejo, R. V. (2017). Comparative Assessment of Goods and Services Provided by Grazing
22 Regulation and Reforestation in Degraded Mediterranean Rangelands. *Land Degradation &*
23 *Development*, 28(4), 1178–1187. <https://doi.org/10.1002/ldr.2368>
- 24 Parsons, S. A., Kutt, A., Vanderduys, E. P., Perry, J. J., & Schwarzkopf, L. (2017). Exploring
25 relationships between native vertebrate biodiversity and grazing land condition. *Rangeland*
26 *Journal*, 39(1), 25–37. <https://doi.org/10.1071/RJ16049>
- 27 Patel, S. (2011). Harmful and beneficial aspects of Parthenium hysterophorus: an update. *3 Biotech*,
28 1(1), 1–9. <https://doi.org/10.1007/s13205-011-0007-7>
- 29 Patten, D. T., Rouse, L., & Stromberg, J. C. (2008). Isolated Spring Wetlands in the Great Basin and
30 Mojave Deserts, USA: Potential Response of Vegetation to Groundwater Withdrawal.
31 *Environmental Management*, 41(3), 398–413. <https://doi.org/10.1007/s00267-007-9035-9>
- 32 Pattison, R. R., Jorgenson, J. C., Reynolds, M. K., & Welker, J. M. (2015). Trends in NDVI and
33 Tundra Community Composition in the Arctic of NE Alaska Between 1984 and 2009.
34 *Ecosystems*, 18(4), 707–719. <https://doi.org/10.1007/s10021-015-9858-9>
- 35 Pedrero, F., Kalavrouziotis, I., Alarcón, J. J., Koukoulakis, P., & Asano, T. (2010). Use of treated
36 municipal wastewater in irrigated agriculture—Review of some practices in Spain and Greece.
37 *Agricultural Water Management*, 97(9), 1233–1241.
38 <https://doi.org/10.1016/J.AGWAT.2010.03.003>
- 39 Pellant, M., Abbey, B., & Karl, S. (2004). Restoring the Great Basin Desert, U.S.A.: Integrating
40 Science, Management, and People. *Environmental Monitoring and Assessment*, 99(1–3), 169–
41 179. <https://doi.org/10.1007/s10661-004-4017-3>
- 42 Pellegrini, A. F. A., Ahlström, A., Hobbie, S. E., Reich, P. B., Nieradzik, L. P., Staver, A. C., ...
43 Jackson, R. B. (2017). Fire frequency drives decadal changes in soil carbon and nitrogen and
44 ecosystem productivity. *Nature*, in press, 3–7. <https://doi.org/10.1038/nature24668>
- 45 Pelletier, J., Paquette, A., Mbindo, K., Zimba, N., Siampale, A., & Chendauka, B. (2018). Carbon sink
46 despite large deforestation in African tropical dry forests (miombo woodlands). *Environmental*
47 *Research Letters*, 13, 094017.

- 1 Pender, J., Mirzabaev, A., & Kato, E. (2009). *Economic analysis of sustainable land*
2 *management options in Central Asia*.
- 3 Peters, D. P., Havstad, K. M., Archer, S. R., & Sala, O. E. (2015). Beyond desertification: new
4 paradigms for dryland landscapes. *Frontiers in Ecology and the Environment*, 13(1), 4–12.
5 <https://doi.org/10.1890/140276>
- 6 Petrie, M. D., Collins, S. L., Swann, A. M., Ford, P. L., & Litvak, M. E. (2015). Grassland to
7 shrubland state transitions enhance carbon sequestration in the northern Chihuahuan Desert.
8 *Global Change Biology*, 21(3), 1226–1235. <https://doi.org/10.1111/gcb.12743>
- 9 Piao, S., Fang, J., Liu, H., & Zhu, B. (2005). NDVI-indicated decline in desertification in China in the
10 past two decades. *Geophysical Research Letters*, 32(6), L06402.
11 <https://doi.org/10.1029/2004GL021764>
- 12 Piao, S., Yin, G., Tan, J., Cheng, L., Huang, M., Li, Y., ... Wang, Y. (2015). Detection and attribution
13 of vegetation greening trend in China over the last 30 years. *Global Change Biology*, 21(4),
14 1601–1609. <https://doi.org/10.1111/gcb.12795>
- 15 Pierre, C., Kergoat, L., Hiernaux, P., Baron, C., Bergametti, G., Rajot, J.-L., ... Marticorena, B.
16 (2017). Impact of Agropastoral Management on Wind Erosion in Sahelian Croplands. *Land*
17 *Degradation & Development*, 29(3), 800–811. <https://doi.org/10.1002/ldr.2783>
- 18 Pierson, F. B., Jason Williams, C., Hardegree, S. P., Clark, P. E., Kormos, P. R., & Al-Hamdan, O. Z.
19 (2013). Hydrologic and Erosion Responses of Sagebrush Steppe Following Juniper
20 Encroachment, Wildfire, and Tree Cutting. *Rangeland Ecology & Management*, 66(3), 274–289.
21 <https://doi.org/10.2111/REM-D-12-00104.1>
- 22 Pierson, F. B., Jason Williams, C., Hardegree, S. P., Weltz, M. A., Stone, J. J., & Clark, P. E. (2011).
23 Fire, Plant Invasions, and Erosion Events on Western Rangelands. *Rangeland Ecology &*
24 *Management*, 64(5), 439–449. <https://doi.org/10.2111/REM-D-09-00147.1>
- 25 Pierzynski, G., Brajendra, Caon, L., & Vargas, R. (2017). Threats to soils: Global trends and
26 perspectives. *Global Land Outlook Working Paper*, 28. Retrieved from
27 [https://static1.squarespace.com/static/5694c48bd82d5e9597570999/t/5931752920099eabdb9b6a](https://static1.squarespace.com/static/5694c48bd82d5e9597570999/t/5931752920099eabdb9b6a7a/1496413492935/Threats+to+Soils__Pierzynski_Brajendra.pdf)
28 [7a/1496413492935/Threats+to+Soils__Pierzynski_Brajendra.pdf](https://static1.squarespace.com/static/5694c48bd82d5e9597570999/t/5931752920099eabdb9b6a7a/1496413492935/Threats+to+Soils__Pierzynski_Brajendra.pdf)
- 29 Piguet, E. (2010). Linking climate change, environmental degradation, and migration: a
30 methodological overview. *Wiley Interdisciplinary Reviews: Climate Change*, 1(4), 517–524.
31 <https://doi.org/10.1002/wcc.54>
- 32 Pilliod, D. S., Welty, J. L., & Arkle, R. S. (2017). Refining the cheatgrass-fire cycle in the Great
33 Basin: Precipitation timing and fine fuel composition predict wildfire trends. *Ecology and*
34 *Evolution*, 7(19), 8126–8151. <https://doi.org/10.1002/ece3.3414>
- 35 Plaza-Bonilla, D., Arrúe, J. L., Cantero-Martínez, C., Fanlo, R., Iglesias, A., & Álvaro-Fuentes, J.
36 (2015). Carbon management in dryland agricultural systems. A review. *Agronomy for*
37 *Sustainable Development*, 35(4), 1319–1334. <https://doi.org/10.1007/s13593-015-0326-x>
- 38 Plaza, C., Zaccone, C., Sawicka, K., Méndez, A. M., Tarquis, A., Gascó, G., ... Maestre, F. T. (2018).
39 Soil resources and element stocks in drylands to face global issues. *Scientific Reports*, 8(1),
40 13788. <https://doi.org/10.1038/s41598-018-32229-0>
- 41 Poesen, J. (2018). Soil erosion in the Anthropocene: Research needs. *Earth Surface Processes and*
42 *Landforms*, 43(1), 64–84. <https://doi.org/10.1002/esp.4250>
- 43 Pointing, S. B., & Belnap, J. (2012). Microbial colonization and controls in dryland systems. *Nature*
44 *Reviews Microbiology*, 10(8), 551–562. <https://doi.org/10.1038/nrmicro2831>
- 45 Pointing, S. B., & Belnap, J. (2014). Disturbance to desert soil ecosystems contributes to dust-
46 mediated impacts at regional scales, 1659–1667. <https://doi.org/10.1007/s10531-014-0690-x>
- 47 Polade, S. D., Gershunov, A., Cayan, D. R., Dettinger, M. D., & Pierce, D. W. (2017). Precipitation in

- 1 a warming world: Assessing projected hydro-climate changes in California and other
2 Mediterranean climate regions. *Scientific Reports*, 7(1), 1–10. [https://doi.org/10.1038/s41598-](https://doi.org/10.1038/s41598-017-11285-y)
3 017-11285-y
- 4 Polley, H. W., Briske, D. D., Morgan, J. A., Wolter, K., Bailey, D. W., & Brown, J. R. (2013).
5 Climate Change and North American Rangelands: Trends, Projections, and Implications.
6 *Rangeland Ecology & Management*, 66(5), 493–511. [https://doi.org/10.2111/REM-D-12-](https://doi.org/10.2111/REM-D-12-00068.1)
7 00068.1
- 8 Pomposi, C., Giannini, A., Kushnir, Y., & Lee, D. E. (2016). Understanding Pacific Ocean influence
9 on interannual precipitation variability in the Sahel. *Geophysical Research Letters*, 43(17),
10 9234–9242. <https://doi.org/10.1002/2016GL069980>
- 11 Pontifes, P. A., García-Meneses, P. M., Gómez-Aíza, L., Monterroso-Rivas, A. I., & Caso Chávez, M.
12 (2018). Land use/land cover change and extreme climatic events in the arid and semi-arid
13 ecoregions of Mexico. *ATMÓSFERA*, 31(4), 355–372.
14 <https://doi.org/10.20937/ATM.2018.31.04.04>
- 15 Popp, A., Domptail, S., Blaum, N., & Jeltsch, F. (2009). Landuse experience does qualify for
16 adaptation to climate change. *Ecological Modelling*, 220(5), 694–702.
17 <https://doi.org/10.1016/J.ECOLMODEL.2008.11.015>
- 18 Porensky, L. M., Mueller, K. E., Augustine, D. J., & Derner, J. D. (2016). Thresholds and gradients in
19 a semi-arid grassland: long-term grazing treatments induce slow, continuous and reversible
20 vegetation change. *Journal of Applied Ecology*, 53(4), 1013–1022. [https://doi.org/10.1111/1365-](https://doi.org/10.1111/1365-2664.12630)
21 2664.12630
- 22 Potter, C., & Weigand, J. (2016). Analysis of Desert Sand Dune Migration Patterns from Landsat
23 Image Time Series for The Southern California Desert. *Journal of Remote Sensing & GIS*, 5(2),
24 1–8. <https://doi.org/10.4172/2469-4134.1000164>
- 25 Poulter, B., Frank, D., Ciais, P., Myneni, R. B., Andela, N., Bi, J., ... van der Werf, G. R. (2014).
26 Contribution of semi-arid ecosystems to interannual variability of the global carbon cycle.
27 *Nature*, 509(7502), 600–603. <https://doi.org/10.1038/nature13376>
- 28 Pourreza, M., Hosseini, S. M., Safari Sinegani, A. A., Matinzadeh, M., & Dick, W. A. (2014). Soil
29 microbial activity in response to fire severity in Zagros oak (*Quercus brantii* Lindl.) forests, Iran,
30 after one year. *Geoderma*, 213, 95–102. <https://doi.org/10.1016/j.geoderma.2013.07.024>
- 31 Powell, M. J. (2009). *Restoration of degraded subtropical thickets in the Baviaanskloof MegaReserve*,
32 *South Africa*. Master Thesis, Rhodes University.
- 33 Pozzi, W., Sheffield, J., Stefanski, R., Cripe, D., Pulwarty, R., Vogt, J. V., ... Lawford, R. (2013).
34 Toward Global Drought Early Warning Capability: Expanding International Cooperation for the
35 Development of a Framework for Monitoring and Forecasting. *Bulletin of the American*
36 *Meteorological Society*, 94(6), 776–785. <https://doi.org/10.1175/BAMS-D-11-00176.1>
- 37 Pradhan, P., Costa, L., Rybski, D., Lucht, W., & Kropp, J. P. (2017). A Systematic Study of
38 Sustainable Development Goal (SDG) Interactions. *Earth's Future*, 5(11), 1169–1179.
39 <https://doi.org/10.1002/2017EF000632>
- 40 Práválie, R. (2016). Drylands extent and environmental issues. A global approach. *Earth-Science*
41 *Reviews*, 161, 259–278. <https://doi.org/10.1016/J.EARSCIREV.2016.08.003>
- 42 Práválie, R., Patriche, C., & Bandoc, G. (2017). Quantification of land degradation sensitivity areas in
43 Southern and Central Southeastern Europe. New results based on improving DISMED
44 methodology with new climate data. *Catena*, 158(May), 309–320.
45 <https://doi.org/10.1016/j.catena.2017.07.006>
- 46 Pricope, N. G., Husak, G., Lopez-Carr, D., Funk, C., & Michaelsen, J. (2013). The climate-population
47 nexus in the East African Horn: Emerging degradation trends in rangeland and pastoral
48 livelihood zones. *Global Environmental Change*, 23(6), 1525–1541.

- 1 <https://doi.org/10.1016/J.GLOENVCHA.2013.10.002>
- 2 Prince, S. D. (2016). Where Does Desertification Occur? Mapping Dryland Degradation at Regional
3 to Global Scales (pp. 225–263). Springer, Berlin, Heidelberg. [https://doi.org/10.1007/978-3-642-](https://doi.org/10.1007/978-3-642-16014-1_9)
4 [16014-1_9](https://doi.org/10.1007/978-3-642-16014-1_9)
- 5 Prince, S. D., De Colstoun, E. B., & Kravitz, L. L. (1998). Evidence from rain-use efficiencies does
6 not indicate extensive Sahelian desertification. *Global Change Biology*, *4*(4), 359–374.
7 <https://doi.org/10.1046/j.1365-2486.1998.00158.x>
- 8 Prince, S., Von Maltitz, G., Zhang, F., Byrne, K., Driscoll, C., Eshel, G., ... Thwin, S. (2018).
9 Chapter 4: Status and trends of land degradation and restoration and associated changes in
10 biodiversity and ecosystem functions. In L. Montanarella, R. Scholes, & A. Brainich (Eds.),
11 *IPBES (2018): The IPBES assessment report on land degradation and restoration*. (pp. 221–
12 338). Bonn, Germany: Secretariat of the Intergovernmental Science-Policy Platform on
13 Biodiversity and Ecosystem Services.
- 14 Prospero, J. M., Blades, E., Mathison, G., & Naidu, R. (2005). Interhemispheric transport of viable
15 fungi and bacteria from Africa to the Caribbean with soil dust. *Aerobiologia*, *21*(1), 1–19.
16 <https://doi.org/10.1007/s10453-004-5872-7>
- 17 Pu, B., & Ginoux, P. (2017). Projection of American dustiness in the late 21st century due to climate
18 change. *Scientific Reports*, *7*(1), 1–10. <https://doi.org/10.1038/s41598-017-05431-9>
- 19 Pulatov, A., Egamberdiev, O., Karimov, A., Tursunov, M., Kienzler, S., Sayre, K., ... Martius, C.
20 (2012). Introducing Conservation Agriculture on Irrigated Meadow Alluvial Soils (Arenosols) in
21 Khorezm, Uzbekistan. In *Cotton, Water, Salts and Soums* (pp. 195–217). Dordrecht: Springer
22 Netherlands. https://doi.org/10.1007/978-94-007-1963-7_13
- 23 Qadir, M., Ghafoor, A., & Murtaza, G. (2000). Amelioration strategies for saline soils: a review. *Land*
24 *Degradation & Development*, *11*(6), 501–521. [https://doi.org/10.1002/1099-](https://doi.org/10.1002/1099-145X(200011/12)11:6<501::AID-LDR405>3.0.CO;2-S)
25 [145X\(200011/12\)11:6<501::AID-LDR405>3.0.CO;2-S](https://doi.org/10.1002/1099-145X(200011/12)11:6<501::AID-LDR405>3.0.CO;2-S)
- 26 Qadir, M., Noble, A. D., Qureshi, A. S., Gupta, R. K., Yuldashev, T., & Karimov, A. (2009). Salt-
27 induced land and water degradation in the Aral Sea basin: A challenge to sustainable agriculture
28 in Central Asia. *Natural Resources Forum*, *33*(2), 134–149. [https://doi.org/10.1111/j.1477-](https://doi.org/10.1111/j.1477-8947.2009.01217.x)
29 [8947.2009.01217.x](https://doi.org/10.1111/j.1477-8947.2009.01217.x)
- 30 Qian, Y., Yasunari, T. J., Doherty, S. J., Flanner, M. G., Lau, W. K. M., Ming, J., ... Zhang, R.
31 (2015). Light-absorbing particles in snow and ice: Measurement and modeling of climatic and
32 hydrological impact. *Advances in Atmospheric Sciences*, *32*(1), 64–91.
33 <https://doi.org/10.1007/s00376-014-0010-0>
- 34 Qin, H., & Liao, T. F. (2016). Labor out-migration and agricultural change in rural China: A
35 systematic review and meta-analysis. *Journal of Rural Studies*, *47*, 533–541.
36 <https://doi.org/10.1016/J.JRURSTUD.2016.06.020>
- 37 Qishlaqi, A., Moore, F., & Forghani, G. (2008). Impact of untreated wastewater irrigation on soils and
38 crops in Shiraz suburban area, SW Iran. *Environmental Monitoring and Assessment*, *141*(1–3),
39 257–273. <https://doi.org/10.1007/s10661-007-9893-x>
- 40 Qureshi, H., Arshad, M., & Bibi, Y. (2014). Invasive flora of Pakistan: a critical analysis.
41 *International Journal of Biosciences (IJB)*, *6655*, 407–424. [https://doi.org/10.12692/ijb/4.1.407-](https://doi.org/10.12692/ijb/4.1.407-424)
42 [424](https://doi.org/10.12692/ijb/4.1.407-424)
- 43 Rachal, D. M., Monger, H. C., Okin, G. S., & Peters, D. C. (2012). Landform influences on the
44 resistance of grasslands to shrub encroachment, Northern Chihuahuan Desert, USA. *Journal of*
45 *Maps*, *8*(4), 507–513. <https://doi.org/10.1080/17445647.2012.727593>
- 46 Racine, C. K. (2008). Soil in the environment: Crucible of terrestrial life: By Daniel Hillel. *Integrated*
47 *Environmental Assessment and Management*, *4*(4), 526.
- 48 Rahman, M. S. (2013). Climate Change, Disaster and Gender Vulnerability: A Study on Two

- 1 Divisions of Bangladesh. *American Journal of Human Ecology*, 2(2), 72–82.
2 <https://doi.org/10.11634/216796221504315>
- 3 Rajot, J. ., Alfaro, S. ., Gomes, L., & Gaudichet, A. (2003). Soil crusting on sandy soils and its
4 influence on wind erosion. *CATENA*, 53(1), 1–16. <https://doi.org/10.1016/S0341->
5 8162(02)00201-1
- 6 Ramankutty, N., Evan, A. T., Monfreda, C., & Foley, J. A. (2008). Farming the planet: 1. Geographic
7 distribution of global agricultural lands in the year 2000. *Global Biogeochemical Cycles*, 22(1),
8 n/a-n/a. <https://doi.org/10.1029/2007GB002952>
- 9 Ramarao, M. V. S., Krishnan, R., Sanjay, J., & Sabin, T. P. (2015). Understanding land surface
10 response to changing South Asian monsoon in a warming climate. *Earth System Dynamics*, 6(2),
11 569–582. <https://doi.org/10.5194/esd-6-569-2015>
- 12 Ramarao, M. V. S., Sanjay, J., Krishnan, R., Mujumdar, M., Bazaz, A., & Revi, A. (2018). On
13 observed aridity changes over the semiarid regions of India in a warming climate. *Theoretical*
14 *and Applied Climatology*, 1–10. <https://doi.org/10.1007/s00704-018-2513-6>
- 15 Rammal, M. M., & Jubair, A. A. (2015). Sand Dunes Stabilization Using Silica Gel and Cement kiln
16 Dust. *Al-Nahrain Journal for Engineering Sciences*, 18(2), 179–191. Retrieved from
17 <https://nahje.com/index.php/main/article/view/176>
- 18 Rao, A. S., Singh, K. C., & Wight, J. R. (1996). Productivity of *Cenchrus Ciliaris* in Relation to Rain-
19 Fall and Fertilization. *Journal of Range Management*, 49(2), 143.
20 <https://doi.org/10.2307/4002684>
- 21 Rao, C. H. H. (2000). Watershed development in India: Recent experience and emerging issues.
22 *Economic and Political Weekly*, 3943–3947.
- 23 Rao, F., Spoor, M., Ma, X., & Shi, X. (2016). Land tenure (in)security and crop-tree intercropping in
24 rural Xinjiang, China. *Land Use Policy*, 50, 102–114.
25 <https://doi.org/10.1016/J.LANDUSEPOL.2015.09.001>
- 26 Rashid, M., Abbas, S. H., & Rehman, A. (2014). the Status of Highly Alien Invasive Plants in
27 Pakistan and Their Impact on the Ecosystem : a Review, 2(1), 2–5.
- 28 Rashki, A., Kaskaoutis, D. G., Rautenbach, C. J. deW., Eriksson, P. G., Qiang, M., & Gupta, P.
29 (2012). Dust storms and their horizontal dust loading in the Sistan region, Iran. *Aeolian*
30 *Research*, 5, 51–62. <https://doi.org/10.1016/J.AEOLIA.2011.12.001>
- 31 Rasmussen, K., D'haen, S., Fensholt, R., Fog, B., Horion, S., Nielsen, J. O., ... Reenberg, A. (2016).
32 Environmental change in the Sahel: reconciling contrasting evidence and interpretations.
33 *Regional Environmental Change*, 16(3), 673–680. <https://doi.org/10.1007/s10113-015-0778-1>
- 34 Rass, N. (2006). *Policies and Strategies to address the vulnerability of pastoralists in Sub-Saharan*
35 *Africa* (PPLPI Working Paper No. 37). Rome. Retrieved from
36 <https://assets.publishing.service.gov.uk/media/57a08c1640f0b64974000fb2/PPLPIwp37.pdf>
- 37 Ratna Reddy, V., Gopinath Reddy, M., Galab, S., Soussan, J., & Springate-Baginski, O. (2004).
38 Participatory Watershed Development in India: Can it Sustain Rural Livelihoods? *Development*
39 *and Change*, 35(2), 297–326. <https://doi.org/10.1111/j.1467-7660.2004.00353.x>
- 40 Ravi, S., Breshears, D. D. D., Huxman, T. T. E., D'Odorico, P., Geomorphology, P. D.-, & 2010, U.
41 (2010). Land degradation in drylands: Interactions among hydrologic–aeolian erosion and
42 vegetation dynamics. *Geomorphology*, 116(3–4), 236–245.
43 <https://doi.org/10.1016/J.GEOMORPH.2009.11.023>
- 44 Reardon, T., B.Barrett, C., A.Berdegúé, J., F.M.Swinnen, J., Barrett, C. B., Berdegúé, J. A., &
45 Swinnen, J. F. M. (2009). Agrifood Industry Transformation and Small Farmers in Developing
46 Countries. *World Development*, 37(11), 1717–1727.
47 <https://doi.org/10.1016/J.WORLDDEV.2008.08.023>

- 1 Reardon, T., Taylor, J. E., Stamoulis, K., Lanjouw, P., & Balisacan, A. (2008). Effects of Non-Farm
2 Employment on Rural Income Inequality in Developing Countries: An Investment Perspective.
3 *Journal of Agricultural Economics*, 51(2), 266–288. [https://doi.org/10.1111/j.1477-](https://doi.org/10.1111/j.1477-9552.2000.tb01228.x)
4 9552.2000.tb01228.x
- 5 Reed, M. S., Buenemann, M., Athlopheng, J., Akhtar-Schuster, M., Bachmann, F., Bastin, G., ...
6 Verzandvoort, S. (2011). Cross-scale monitoring and assessment of land degradation and
7 sustainable land management: A methodological framework for knowledge management. *Land*
8 *Degradation & Development*, 22(2), 261–271. <https://doi.org/10.1002/ldr.1087>
- 9 Reed, M. S., Nkonya, E., Winslow, M., Mortimore, M., & Mirzabaev, A. (2011). Monitoring and
10 assessing the influence of social, economic and policy factors on sustainable land management
11 in drylands. *Land Degradation & Development*, 22(2), 240–247.
12 <https://doi.org/10.1002/ldr.1048>
- 13 Reed, M. S., Stringer, L. C., Dougill, A. J., Perkins, J. S., Athlopheng, J. R., Mulale, K., & Favretto,
14 N. (2015). Reorienting land degradation towards sustainable land management: Linking
15 sustainable livelihoods with ecosystem services in rangeland systems. *Journal of Environmental*
16 *Management*, 151, 472–485. <https://doi.org/10.1016/J.JENVMAN.2014.11.010>
- 17 Reichardt, K. (2010). College on Soil Physics : Soil Physical Properties and Processes under Climate
18 Change. <https://doi.org/10.1016/j.still.2004.07.002>
- 19 Reicosky, D., & Crovetto, C. (2014). No-till systems on the Chequen Farm in Chile: A success story
20 in bringing practice and science together. *International Soil and Water Conservation Research*,
21 2(1), 66–77. [https://doi.org/10.1016/S2095-6339\(15\)30014-9](https://doi.org/10.1016/S2095-6339(15)30014-9)
- 22 Reid, H., Bourne, A., Muller, H., Podvin, K., Scorgie, S., & Orindi, V. (2018). Chapter 16 - A
23 Framework for Assessing the Effectiveness of Ecosystem-Based Approaches to Adaptation. In
24 Z. Zommers & K. B. T.-R. Alverson (Eds.), *Resilience - The Science of Adaptation to Climate*
25 *Change* (pp. 207–216). London, UK: Elsevier. [https://doi.org/https://doi.org/10.1016/B978-0-](https://doi.org/https://doi.org/10.1016/B978-0-12-811891-7.00016-5)
26 12-811891-7.00016-5
- 27 Reij, C., & Waters-Bayer, A. (Eds.). (2001). *Farmer innovation in Africa: A source of inspiration for*
28 *agricultural development*. London, UK: Earthscan Publications Ltd.
- 29 Rengasamy, P. (2006). World salinization with emphasis on Australia. *Journal of Experimental*
30 *Botany*, 57(5), 1017–1023. <https://doi.org/10.1093/jxb/erj108>
- 31 Réquier-Desjardins M, & Bied-Charreton, M. (2006). *Évaluation économique des coûts économiques*
32 *et sociaux de la désertification en Afrique*,. Paris.
- 33 Revi, A., & Rosenzweig, C. (2013). *The Urban Opportunity to Enable Transformative and*
34 *Sustainable development. Background paper for the High-Level Panel of Eminent Persons on*
35 *the Post-2015 Development Agenda. Prepared by the co-chairs of the Sustainable Development*
36 *Solutions Network Themat.*
- 37 Rey, A., Pegoraro, E., Oyonarte, C., Were, A., Escribano, P., & Raimundo, J. (2011). Impact of land
38 degradation on soil respiration in a steppe (*Stipa tenacissima* L.) semi-arid ecosystem in the SE
39 of Spain. *Soil Biology and Biochemistry*, 43(2), 393–403.
40 <https://doi.org/10.1016/j.soilbio.2010.11.007>
- 41 Rey, B., Fuller, A., Mitchell, D., Meyer, L. C. R., & Hetem, R. S. (2017). Drought-induced starvation
42 of aardvarks in the Kalahari: an indirect effect of climate change. *Biology Letters*, 13(7),
43 20170301. <https://doi.org/10.1098/rsbl.2017.0301>
- 44 Reyer, C. P. ., Otto, I. M., Adams, S., Albrecht, T., Baarsch, F., Carlsburg, M., ... Stagl, J. (2017).
45 Climate change impacts in Central Asia and their implications for development. *Regional*
46 *Environmental Change*, 17(6), 1639–1650. <https://doi.org/10.1007/s10113-015-0893-z>
- 47 Reyer, C. P. O., Leuzinger, S., Rammig, A., Wolf, A., Bartholomeus, R. P., Bonfante, A., ... Pereira,
48 M. (2013). A plant's perspective of extremes: terrestrial plant responses to changing climatic

- 1 variability. *Global Change Biology*, 19(1), 75–89. <https://doi.org/10.1111/gcb.12023>
- 2 Reyes-García, V., Guèze, M., Luz, A. C., Paneque-Gálvez, J., Macía, M. J., Orta-Martínez, M., ...
3 Rubio-Campillo, X. (2013). Evidence of traditional knowledge loss among a contemporary
4 indigenous society. *Evolution and Human Behavior*, 34(4), 249–257.
5 <https://doi.org/10.1016/J.EVOLHUMBEHAV.2013.03.002>
- 6 Reynolds, J. F., Grainger, A., Stafford Smith, D. M., Bastin, G., Garcia-Barrios, L., Fernández, R.
7 J., ... Zdruli, P. (2011). Scientific concepts for an integrated analysis of desertification. *Land*
8 *Degradation & Development*, 22(2), 166–183. <https://doi.org/10.1002/ldr.1104>
- 9 Reynolds, J. F., Kemp, P. R., Ogle, K., & Fernández, R. J. (2004). Modifying the ‘pulse–reserve’
10 paradigm for deserts of North America: precipitation pulses, soil water, and plant responses.
11 *Oecologia*, 141(2), 194–210. <https://doi.org/10.1007/s00442-004-1524-4>
- 12 Reynolds, J. F., Smith, D. M. S., Lambin, E. F., Turner, B. L., Mortimore, M., Batterbury, S. P. J., ...
13 Walker, B. (2007). Global Desertification: Building a Science for Dryland Development.
14 *Science*, 316(5826), 847–851. <https://doi.org/10.1126/science.1131634>
- 15 Richardson, D. M., Carruthers, J., Hui, C., Impson, F. A. C., Miller, J. T., Robertson, M. P., ...
16 Wilson, J. R. U. (2011). Human-mediated introductions of Australian acacias - a global
17 experiment in biogeography. *Diversity and Distributions*, 17(5), 771–787.
18 <https://doi.org/10.1111/j.1472-4642.2011.00824.x>
- 19 Rishmawi, K., Prince, S., Rishmawi, K., & Prince, S. D. (2016). Environmental and Anthropogenic
20 Degradation of Vegetation in the Sahel from 1982 to 2006. *Remote Sensing*, 8(11), 948.
21 <https://doi.org/10.3390/rs8110948>
- 22 Robinson, B. E., Masuda, Y. J., Kelly, A., Holland, M. B., Bedford, C., Childress, M., ... Veit, P.
23 (2018). Incorporating Land Tenure Security into Conservation. *Conservation Letters*, 11(2),
24 e12383. <https://doi.org/10.1111/conl.12383>
- 25 Roderick, M. L., Greve, P., & Farquhar, G. D. (2015). On the assessment of aridity with changes in
26 atmospheric CO₂. *Water Resources Research*, 51(7), 5450–5463.
27 <https://doi.org/10.1002/2015WR017031>
- 28 Rodima-Taylor, D., Olwig, M. F., & Chhetri, N. (2012). Adaptation as innovation, innovation as
29 adaptation: An institutional approach to climate change. *Applied Geography*, 33, 107–111.
30 <https://doi.org/10.1016/J.APGEOG.2011.10.011>
- 31 Rodriguez-Caballero, E., Belnap, J., Büdel, B., Crutzen, P. J., Andreae, M. O., Pöschl, U., & Weber,
32 B. (2018). Dryland photoautotrophic soil surface communities endangered by global change.
33 *Nature Geoscience*, 11(3), 185–189. <https://doi.org/10.1038/s41561-018-0072-1>
- 34 Rojas-Downing, M. M., Nejadhashemi, A. P., Harrigan, T., & Woznicki, S. A. (2017). Climate
35 change and livestock: Impacts, adaptation, and mitigation. *Climate Risk Management*, 16, 145–
36 163. <https://doi.org/10.1016/j.crm.2017.02.001>
- 37 Rosenfeld, D., Rudich, Y., & Lahav, R. (2001). Desert dust suppressing precipitation: a possible
38 desertification feedback loop. *Proceedings of the National Academy of Sciences of the United*
39 *States of America*, 98(11), 5975–80. <https://doi.org/10.1073/pnas.101122798>
- 40 Rosenzweig, C., Elliott, J., Deryng, D., Ruane, A. C., Müller, C., Arneth, A., ... Jones, J. W. (2014).
41 Assessing agricultural risks of climate change in the 21st century in a global gridded crop model
42 intercomparison. *Proceedings of the National Academy of Sciences of the United States of*
43 *America*, 111(9), 3268–73. <https://doi.org/10.1073/pnas.1222463110>
- 44 Rotenberg, E., & Yakir, D. (2010). Contribution of semi-arid forests to the climate system. *Science*
45 *(New York, N.Y.)*, 327(5964), 451–4. <https://doi.org/10.1126/science.1179998>
- 46 Rotstayn, L. D., & Lohmann, U. (2002). Tropical rainfall trends and the indirect aerosol effect.
47 *Journal of Climate*, 15(15), 2103–2116. [https://doi.org/10.1175/1520-0442\(2002\)015<2103:TRTATI>2.0.CO;2](https://doi.org/10.1175/1520-0442(2002)015<2103:TRTATI>2.0.CO;2)

- 1 Rouget, M., Robertson, M. P., Wilson, J. R. U., Hui, C., Essl, F., Renteria, J. L., & Richardson, D. M.
2 (2016). Invasion debt - quantifying future biological invasions. *Diversity and Distributions*,
3 22(4), 445–456. <https://doi.org/10.1111/ddi.12408>
- 4 Ruiz, J., Bielza, M., Garrido, A., & Iglesias, A. (2015). Dealing with drought in irrigated agriculture
5 through insurance schemes: an application to an irrigation district in Southern Spain. *Spanish*
6 *Journal of Agricultural Research*, 13(4), e0106. <https://doi.org/10.5424/sjar/2015134-6941>
- 7 Rusike, J., Twomlow, S., Freeman, H. A., & Heinrich, G. M. (2006). Does farmer participatory
8 research matter for improved soil fertility technology development and dissemination in
9 Southern Africa? *International Journal of Agricultural Sustainability*, 4(3), 176–192.
10 <https://doi.org/10.1080/14735903.2006.9684801>
- 11 Rutherford, W. A., Painter, T. H., Ferrenberg, S., Belnap, J., Okin, G. S., Flagg, C., & Reed, S. C.
12 (2017). Albedo feedbacks to future climate via climate change impacts on dryland biocrusts.
13 *Scientific Reports*, 7(1), 44188. <https://doi.org/10.1038/srep44188>
- 14 Rutledge, S., Campbell, D. I., Baldocchi, D., & Schipper, L. A. (2010). Photodegradation leads to
15 increased carbon dioxide losses from terrestrial organic matter. *Global Change Biology*, 16(11),
16 3065–3074. <https://doi.org/10.1111/j.1365-2486.2009.02149.x>
- 17 Sacande, M. (2018). *Action Against Desertification, Land Restoration*. Rome, Italy. Retrieved from
18 <http://www.fao.org/3/i9144en/I9144EN.pdf>
- 19 Safriel, U., Adeel, Z., Niemeijer, D., Puigdefabregas, J., White, R., Lal, R., Winslow, M., ... McNab,
20 D. (2005). Dryland systems. In R. Hassan, R. J. Scholes, & N. Ash (Eds.), *Ecosystems and*
21 *Human Well-Being: Current State and Trends* (The Millen, pp. 623–662). Washington DC:
22 Island Press.
- 23 Safriel, U. (2009). Deserts and desertification: Challenges but also opportunities. *Land Degradation*
24 *& Development*, 20(4), 353–366. <https://doi.org/10.1002/ldr.935>
- 25 Safriel, U., & Adeel, Z. (2008). Development paths of drylands: thresholds and sustainability.
26 *Sustainability Science*, 3(1), 117–123. <https://doi.org/10.1007/s11625-007-0038-5>
- 27 Safriel, U. N. (2007). The Assessment of Global Trends in Land Degradation. In *Climate and Land*
28 *Degradation* (pp. 1–38). Berlin, Heidelberg: Springer Berlin Heidelberg.
29 https://doi.org/10.1007/978-3-540-72438-4_1
- 30 Safronov, O., Kreuzwieser, J., Haberer, G., Alyousif, M. S., Schulze, W., Al-Harbi, N., ...
31 Kangasjärvi, J. (2017). Detecting early signs of heat and drought stress in Phoenix dactylifera
32 (date palm). *PLOS ONE*, 12(6), e0177883. <https://doi.org/10.1371/journal.pone.0177883>
- 33 Saha, M. V., Scanlon, T. M., & D’Odorico, P. (2015). Examining the linkage between shrub
34 encroachment and recent greening in water-limited southern Africa. *Ecosphere*, 6(9), art156.
35 <https://doi.org/10.1890/ES15-00098.1>
- 36 Sahara and Sahel Observatory (OSS). (2016). The Great Green Wall | A Development Programme for
37 the Sahara and the Sahel: Projects monitoring and evaluation approach based on geospatial
38 applications | Sahara and Sahel Observatory. Retrieved May 25, 2018, from [http://www.oss-](http://www.oss-online.org/en/great-green-wall-development-programme-sahara-and-sahel-projects-monitoring-and-evaluation-approach)
39 [online.org/en/great-green-wall-development-programme-sahara-and-sahel-projects-monitoring-](http://www.oss-online.org/en/great-green-wall-development-programme-sahara-and-sahel-projects-monitoring-and-evaluation-approach)
40 [and-evaluation-approach](http://www.oss-online.org/en/great-green-wall-development-programme-sahara-and-sahel-projects-monitoring-and-evaluation-approach)
- 41 Saidan, M., Albaali, A. G., Alasis, E., & Kaldellis, J. K. (2016). Experimental study on the effect of
42 dust deposition on solar photovoltaic panels in desert environment. *Renewable Energy*, 92, 499–
43 505. <https://doi.org/10.1016/J.RENENE.2016.02.031>
- 44 Saifi, M., Boulghobra, N. and Fattoum, L., & Oesterheld, M. (2015). The Green Dam in Algeria as a
45 tool to combat desertification. *Planet@Risk*, 3(1), 68–71.
46 <https://doi.org/10.1016/j.jaridenv.2006.01.021>
- 47 Sakai, Y., Ma, Y., Xu, C., Wu, H., Zhu, W., & Yang, J. (2012). Phytodesalination of a Salt-Affected
48 Soil with Four Halophytes in China. *J. Arid Land Studies*, 22(1), 239–302. Retrieved from

- 1 http://nodaiweb.university.jp/desert/pdf/JALS-B02_17-20.pdf
- 2 Sala, A., Verdaguer, D., & Vila, M. (2006). Sensitivity of the Invasive Geophyte *Oxalis pes-caprae* to
3 Nutrient Availability and Competition. *Annals of Botany*, 99(4), 637–645.
4 <https://doi.org/10.1093/aob/mcl289>
- 5 Salamani, M., Hanifi, H. K., Hirche, A., & Nedjraoui, D. (2012). Évaluation de la sensibilité À la
6 désertification en algÉrie. *Rev. Écol. (Terre Vie)*, 67. Retrieved from
7 [http://documents.irevues.inist.fr/bitstream/handle/2042/55954/RevuedEcologie_2013_68_1_71.](http://documents.irevues.inist.fr/bitstream/handle/2042/55954/RevuedEcologie_2013_68_1_71.pdf?sequence=1)
8 [pdf?sequence=1](http://documents.irevues.inist.fr/bitstream/handle/2042/55954/RevuedEcologie_2013_68_1_71.pdf?sequence=1)
- 9 Salami, H., Shahnooshi, N., & Thomson, K. J. (2009). The economic impacts of drought on the
10 economy of Iran: An integration of linear programming and macroeconomic modelling
11 approaches. *Ecological Economics*, 68(4), 1032–1039.
12 <https://doi.org/10.1016/J.ECOLECON.2008.12.003>
- 13 Salik, K. M., Qaisrani, A., Umar, M. A., & Ali, S. M. (2017). *Migration futures in Asia and Africa:*
14 *economic opportunities and distributional effects – the case of Pakistan*. Sustainable
15 Development Policy Institute. Retrieved from <https://think-asia.org/handle/11540/7572>
- 16 Salvati, L. (2014). A socioeconomic profile of vulnerable land to desertification in Italy. *Science of*
17 *the Total Environment*, 466–467, 287–299. Retrieved from
18 <http://www.sciencedirect.com/science/article/pii/S0048969713007432>
- 19 Salvati, L., Zitti, M., Ceccarelli, T., & Perini, L. (2009). Developing a Synthetic Index of Land
20 Vulnerability to Drought and Desertification. *Geographical Research*, 47(3), 280–291.
21 <https://doi.org/10.1111/j.1745-5871.2009.00590.x>
- 22 Samoli, E., Kougea, E., Kassomenos, P., Analitis, A., & Katsouyanni, K. (2011). Does the presence of
23 desert dust modify the effect of PM10 on mortality in Athens, Greece? *Science of The Total*
24 *Environment*, 409(11), 2049–2054. <https://doi.org/10.1016/J.SCITOTENV.2011.02.031>
- 25 Samuel, A., Joy, K. ., Paranjape, S., Peddi, S., Adagale, R., Deshpande, P., & Kulkarni, S. (2007).
26 *Watershed development in Maharashtra: present scenario and issues for restructuring the*
27 *programme*.
- 28 Samuel, A., Joy, K. J., & Joy, K. J. (2018). Changing land use, agrarian context and rural
29 transformation. In *India's Water Futures* (pp. 57–78). Routledge India.
30 <https://doi.org/10.4324/9780429423529-4>
- 31 Sanandiya, N. D., & Siddhanta, A. K. (2014). Chemical studies on the polysaccharides of *Salicornia*
32 *brachiata*. *Carbohydrate Polymers*, 112, 300–307.
33 <https://doi.org/10.1016/J.CARPOL.2014.05.072>
- 34 Sanz, M. J., Vente, J. de, Chotte, J.-L., Bernoux, M., Kust, G., Ruiz, I., ... Akhtar-Schuster, M.
35 (2017). *Sustainable Land Management contribution to successful land-based climate change*
36 *adaptation and mitigation. A Report of the Science-Policy Interface*. Bonn, Germany.
- 37 Saparov, A. (2014). Soil Resources of the Republic of Kazakhstan: Current Status, Problems and
38 Solutions (pp. 61–73). Springer, Cham. https://doi.org/10.1007/978-3-319-01017-5_2
- 39 Sarver, T., Al-Qaraghuli, A., & Kazmerski, L. L. (2013). A comprehensive review of the impact of
40 dust on the use of solar energy: History, investigations, results, literature, and mitigation
41 approaches. *Renewable and Sustainable Energy Reviews*, 22, 698–733.
42 <https://doi.org/10.1016/J.RSER.2012.12.065>
- 43 Scasta, J. D., Engle, D. M., Fuhlendorf, S. D., Redfearn, D. D., & Bidwell, T. G. (2015). Meta-
44 Analysis of Exotic Forages as Invasive Plants in Complex Multi-Functioning Landscapes.
45 *Invasive Plant Science and Management*, 8(03), 292–306. [https://doi.org/10.1614/IPSM-D-14-](https://doi.org/10.1614/IPSM-D-14-00076.1)
46 [00076.1](https://doi.org/10.1614/IPSM-D-14-00076.1)
- 47 Scheff, J., & Frierson, D. M. W. (2012). Robust future precipitation declines in CMIP5 largely reflect
48 the poleward expansion of model subtropical dry zones. *Geophysical Research Letters*, 39(18).

- 1 <https://doi.org/10.1029/2012GL052910>
- 2 Scheff, J., & Frierson, D. M. W. W. (2015). Terrestrial aridity and its response to greenhouse
3 warming across CMIP5 climate models. *Journal of Climate*, 28(14), 5583–5600.
4 <https://doi.org/10.1175/JCLI-D-14-00480.1>
- 5 Schewe, J., & Levermann, A. (2017). Non-linear intensification of Sahel rainfall as a possible
6 dynamic response to future warming. *Earth Syst. Dynam.*, 8, 495–505.
7 <https://doi.org/10.5194/esd-8-495-2017>
- 8 Schiappacasse, I., Nahuelhual, L., Vásquez, F., & Echeverría, C. (2012). Assessing the benefits and
9 costs of dryland forest restoration in central Chile. *Journal of Environmental Management*, 97,
10 38–45. Retrieved from <http://www.sciencedirect.com/science/article/pii/S0301479711004178>
- 11 Schilling, J., Akuno, M., Scheffran, J., & Weinzierl, T. (2014). On raids and relations: Climate
12 change, pastoral conflict and adaptation in northwestern Kenya. *Conflict-Sensitive Adaptation to*
13 *Climate Change in Africa*. Retrieved from
14 https://scholar.google.com/scholar?hl=en&as_sdt=0%2C25&q=On+raids+and+relations%3A+C
15 [limite+change%2C+pastoral+conflict+and+adaptation+in+north-western+Kenya&btnG=](https://scholar.google.com/scholar?hl=en&as_sdt=0%2C25&q=On+raids+and+relations%3A+C+climate+change%2C+pastoral+conflict+and+adaptation+in+north-western+Kenya&btnG=)
- 16 Schlenker, W., & Lobell, D. B. (2010). Robust negative impacts of climate change on African
17 agriculture. *Environmental Research Letters*, 5(1), 014010. [https://doi.org/10.1088/1748-](https://doi.org/10.1088/1748-9326/5/1/014010)
18 [9326/5/1/014010](https://doi.org/10.1088/1748-9326/5/1/014010)
- 19 Schlesinger, C., White, S., & Muldoon, S. (2013). Spatial pattern and severity of fire in areas with and
20 without buffel grass (*Cenchrus ciliaris*) and effects on native vegetation in central Australia.
21 *Austral Ecology*, 38(7), 831–840. <https://doi.org/10.1111/aec.12039>
- 22 Schlesinger, P., Mamane, Y., & Grishkan, I. (2006). Transport of microorganisms to Israel during
23 Saharan dust events. *Aerobiologia*, 22(4), 259–273. <https://doi.org/10.1007/s10453-006-9038-7>
- 24 Schlesinger, W. H. H., Reynolds, J. F. F., Cunningham, G. L. L., Hueneke, L. F. F., Jarrell, W. M.
25 M., Virginia, R. A. A., & Whitford, W. G. G. (1990). Biological Feedbacks in Global
26 Desertification. *Science*, 4946(4946), 1043–1048. <https://doi.org/10.1126/science.247.4946.1043>
- 27 Schmidt, E., Chinowsky, P., Robinson, S., & Strzepek, K. (2017). Determinants and impact of
28 sustainable land management (SLM) investments: A systems evaluation in the Blue Nile Basin,
29 Ethiopia. *Agricultural Economics*, 48(5), 613–627. <https://doi.org/10.1111/agec.12361>
- 30 Schneider, F. D., & Kéfi, S. (2016). Spatially heterogeneous pressure raises risk of catastrophic shifts.
31 *Theoretical Ecology*, 9(2), 207–217. <https://doi.org/10.1007/s12080-015-0289-1>
- 32 Schofield, R. V., & Kirkby, M. J. (2003). Application of salinization indicators and initial
33 development of potential global soil salinization scenario under climatic change. *Global*
34 *Biogeochemical Cycles*, 17(3), n/a-n/a. <https://doi.org/10.1029/2002GB001935>
- 35 Scholes, R. J. (2009). Syndromes of dryland degradation in southern Africa. *African Journal of Range*
36 *and Forage Science*, 26(3), 113–125. <https://doi.org/10.2989/AJRF.2009.26.3.2.947>
- 37 Scholes, R. J., & Archer, S. R. (1997). Tree-grass interactions in savannas. *Annual Review of Ecology*
38 *and Systematics*, 28, 517–544.
- 39 Scholes, R. J., & Walker, B. H. (1993). *An African savanna: synthesis of the Nylsvley study*.
40 Cambridge: Cambridge University Press.
- 41 Schooley, R. L., Bestelmeyer, B. T., & Campanella, A. (2018). Shrub encroachment, productivity
42 pulses, and core-transient dynamics of Chihuahuan Desert rodents. *Ecosphere*, 9(7).
43 <https://doi.org/10.1002/ecs2.2330>
- 44 Schulz, J. J., Cayuela, L., Echeverria, C., Salas, J., & Rey Benayas, J. M. (2010). Monitoring land
45 cover change of the dryland forest landscape of Central Chile (1975–2008). *Applied Geography*,
46 30(3), 436–447. <https://doi.org/10.1016/J.APGEOG.2009.12.003>
- 47 Scott-shaw, R., & Morris, C. D. (2015). Grazing depletes forb species diversity in the mesic

- 1 grasslands of KwaZulu-Natal , South Africa. *African Journal of Range & Forage Science*, 32(1),
2 37–41. <https://doi.org/10.2989/10220119.2014.901418>
- 3 Seaquist, J. W., Hickler, T., Eklundh, L., Ardö, J., & Heumann, B. W. (2009). Disentangling the
4 effects of climate and people on Sahel vegetation dynamics. *Biogeosciences*, 6(3), 469–477.
5 <https://doi.org/10.5194/bg-6-469-2009>
- 6 Sebaa A., Berroussi S., Bouhanna M., Boulghobra N., Hadri T., Koull N., Kherraze M., B. T.
7 (2015). *Guide des techniques de lutte contre l'ensablement au Sahara Algérien* (CRSTRA).
8 Retrieved from [https://rs.umc.edu.dz/umc/ouvrage/Guide](https://rs.umc.edu.dz/umc/ouvrage/Guide%20technique%20de%20lutte%20contre%20l'ensablement.pdf) technique de lutte contre
9 l'ensablement.pdf
- 10 Secretariat of the Convention on Biological Diversity. (2010). *Pastoralism, Nature Conservation and*
11 *Development: A Good Practice Guide*. Montreal.
- 12 Seddon, A. W. R., Macias-Fauria, M., Long, P. R., Benz, D., & Willis, K. J. (2016). Sensitivity of
13 global terrestrial ecosystems to climate variability. *Nature*, 531(7593), 229–232.
14 <https://doi.org/10.1038/nature16986>
- 15 Seebens, H., Blackburn, T. M., Dyer, E. E., Genovesi, P., Hulme, P. E., Jeschke, J. M., ... Essl, F.
16 (2018). Global rise in emerging alien species results from increased accessibility of new source
17 pools. *Proceedings of the National Academy of Sciences of the United States of America*,
18 115(10), E2264–E2273. <https://doi.org/10.1073/pnas.1719429115>
- 19 Seebens, H., Essl, F., Dawson, W., Fuentes, N., Moser, D., Pergl, J., ... Blasius, B. (2015). Global
20 trade will accelerate plant invasions in emerging economies under climate change. *Global*
21 *Change Biology*, 21(11), 4128–4140. <https://doi.org/10.1111/gcb.13021>
- 22 Sefelnasr, A., Gossel, W., & Wycisk, P. (2014). Three-dimensional groundwater flow modeling
23 approach for the groundwater management options for the Dakhla oasis, Western Desert, Egypt.
24 *Environmental Earth Sciences*, 72(4), 1227–1241. <https://doi.org/10.1007/s12665-013-3041-4>
- 25 Seidel, D. J., & Randel, W. J. (2007). Recent widening of the tropical belt: Evidence from tropopause
26 observations. *Journal of Geophysical Research*, 112(D20), D20113.
27 <https://doi.org/10.1029/2007JD008861>
- 28 Sen, S., & Ganguly, S. (2017). Opportunities, barriers and issues with renewable energy development
29 – A discussion. *Renewable and Sustainable Energy Reviews*, 69, 1170–1181.
30 <https://doi.org/10.1016/J.RSER.2016.09.137>
- 31 Seneviratne, S. I., & Ciais, P. (2017). Environmental science: Trends in ecosystem recovery from
32 drought. *Nature*, 548(7666), 164–165. <https://doi.org/10.1038/548164a>
- 33 Seneviratne, S. I., Corti, T., Davin, E. L., Hirschi, M., Jaeger, E. B., Lehner, I., ... Teuling, A. J.
34 (2010). Investigating soil moisture–climate interactions in a changing climate: A review. *Earth-*
35 *Science Reviews*, 99(3–4), 125–161. <https://doi.org/10.1016/J.EARSCIREV.2010.02.004>
- 36 Serdeczny, O., Adams, S., Baarsch, F., Coumou, D., Robinson, A., Hare, W., ... Reinhardt, J. (2017).
37 Climate change impacts in Sub-Saharan Africa: from physical changes to their social
38 repercussions. *Regional Environmental Change*, 17(6), 1585–1600.
39 <https://doi.org/10.1007/s10113-015-0910-2>
- 40 Sergeant, C. J., Moynahan, B. J., & Johnson, W. F. (2012). Practical advice for implementing long-
41 term ecosystem monitoring. *Journal of Applied Ecology*, 49(5), 969–973.
42 <https://doi.org/10.1111/j.1365-2664.2012.02149.x>
- 43 Serpa, D., Nunes, J. P., Santos, J., Sampaio, E., Jacinto, R., Veiga, S., ... Abrantes, N. (2015). Impacts
44 of climate and land use changes on the hydrological and erosion processes of two contrasting
45 Mediterranean catchments. *Science of the Total Environment*, 538, 64–77.
46 <https://doi.org/10.1016/j.scitotenv.2015.08.033>
- 47 Serrano-Ortiz, P., Were, A., Reverter, B. R., Villagarcía, L., Domingo, F., Dolman, A. J., &
48 Kowalski, A. S. (2015). Seasonality of net carbon exchanges of Mediterranean ecosystems

- 1 across an altitudinal gradient. *Journal of Arid Environments*, 115, 1–9.
2 <https://doi.org/10.1016/j.jaridenv.2014.12.003>
- 3 Settele, J., R. Scholes, R. Betts, S. Bunn, P. Leadley, D. Nepstad, J.T. Overpeck, and M. A. T.
4 (2014). Terrestrial and inland water systems. In *Climate Change 2014: Impacts, Adaptation, and*
5 *Vulnerability. Part A: Global and Sectoral Aspects*. (pp. 271–359).
- 6 Shabani, F., Kumar, L., & Taylor, S. (2015). Distribution of Date Palms in the Middle East Based on
7 Future Climate Scenarios. *Experimental Agriculture*, 51(02), 244–263.
8 <https://doi.org/10.1017/S001447971400026X>
- 9 Shackleton, C. M., & Shackleton, S. E. (2004). Use of woodland resources for direct household
10 provisioning. In M. J. Lawes, H. A. C. Eeley, C. M. Shackleton, & B. G. S. Geach (Eds.),
11 *Indigenous forests and woodlands in South Africa*. Pietermaritzberg, South Africa: University of
12 KwaZulu-Natal Press.
- 13 Sharkhuu, A., Plante, A. F., Enkhmandal, O., Gonneau, C., Casper, B. B., Boldgiv, B., & Petraitis, P.
14 S. (2016). Soil and ecosystem respiration responses to grazing, watering and experimental
15 warming chamber treatments across topographical gradients in northern Mongolia. *Geoderma*,
16 269, 91–98. <https://doi.org/10.1016/j.geoderma.2016.01.041>
- 17 Sharma, R., Wungrampha, S., Singh, V., Pareek, A., & Sharma, M. K. (2016). Halophytes as
18 bioenergy crops. *Frontiers in Plant Science*, 7, 1372. <https://doi.org/10.3389/fpls.2016.01372>
- 19 Shaw, E. C., Gabric, A. J., & McTainsh, G. H. (2008). Impacts of aeolian dust deposition on
20 phytoplankton dynamics in Queensland coastal waters. *Marine and Freshwater Research*,
21 59(11), 951. <https://doi.org/10.1071/MF08087>
- 22 Sheen, K. L., Smith, D. M., Dunstone, N. J., Eade, R., Rowell, D. P., & Vellinga, M. (2017). Skilful
23 prediction of Sahel summer rainfall on inter-annual and multi-year timescales. *Nature*
24 *Communications*, 8, 14966. <https://doi.org/10.1038/ncomms14966>
- 25 Sheffield, J., & Wood, E. F. (2008). Projected changes in drought occurrence under future global
26 warming from multi-model, multi-scenario, IPCC AR4 simulations. *Climate Dynamics*, 31(1),
27 79–105. <https://doi.org/10.1007/s00382-007-0340-z>
- 28 Sheffield, J., Wood, E. F., & Roderick, M. L. (2012). Little change in global drought over the past 60
29 years. *Nature*, 491(7424), 435–438. <https://doi.org/10.1038/nature11575>
- 30 Sherbinin, A. de, & Bai, L. (2018). Geospatial modeling and mapping, 85–91.
31 <https://doi.org/10.4324/9781315638843-6>
- 32 Sherif, M. M., & Singh, V. P. (1999). Effect of climate change on sea water intrusion in coastal
33 aquifers. *Hydrological Processes*, 13(8), 1277–1287. [https://doi.org/10.1002/\(SICI\)1099-1085\(19990615\)13:8<1277::AID-HYP765>3.0.CO;2-W](https://doi.org/10.1002/(SICI)1099-1085(19990615)13:8<1277::AID-HYP765>3.0.CO;2-W)
- 35 Sherwood, S., & Fu, Q. (2014, February 14). A drier future? *Science*. American Association for the
36 Advancement of Science. <https://doi.org/10.1126/science.1247620>
- 37 Shiferaw, B. A., Okello, J., & Reddy, R. V. (2009). Adoption and adaptation of natural resource
38 management innovations in smallholder agriculture: reflections on key lessons and best
39 practices. *Environment, Development and Sustainability*, 11(3), 601–619.
40 <https://doi.org/10.1007/s10668-007-9132-1>
- 41 Shiferaw, B., & Djido, A. (2016, July 29). Patterns of Labor Productivity and Income Diversification
42 in the Rural Farm and Non-farm Sectors in Sub-Saharan Africa. Partnership for Economic
43 Policy (PEP). Retrieved from <https://www.africaportal.org/publications/patterns-labor-productivity-and-income-diversification-rural-farm-and-non-farm-sectors-sub-saharan-africa/>
- 45 Short, A. D., O'Connor, T. G., & Hurt, C. R. (2003). Medium-term changes in grass composition and
46 diversity of Highland Sourveld grassland in the southern Drakensberg in response to fire and
47 grazing management. *African Journal of Range and Forage Science*, 20(1), 1–10.
48 <https://doi.org/10.2989/10220110309485792>

- 1 Siebert, C., Rödiger, T., Schulz, S., Horovitz, M., Merz, R., Friesen, J., ... Schüth, C. (2016). New
2 Tools for Coherent Information Base for IWRM in Arid Regions: The Upper Mega Aquifer
3 System on the Arabian Peninsula. In *Integrated Water Resources Management: Concept,
4 Research and Implementation* (pp. 85–106). Cham: Springer International Publishing.
5 https://doi.org/10.1007/978-3-319-25071-7_4
- 6 Sietz, D., Lüdeke, M. K. B. M., Change, C. W., Walther, C., & Change, C. W. (2011). Categorisation
7 of typical vulnerability patterns in global drylands. *Global Environmental Change*, *21*(2), 431–
8 440. <https://doi.org/10.1016/J.GLOENVCHA.2010.11.005>
- 9 Sietz, D., Ordoñez, J. C., Kok, M. T. J., Janssen, P., Hilderink, H. B. M., Tittone, P., & Van Dijk, H.
10 (2017). Nested archetypes of vulnerability in African drylands: where lies potential for
11 sustainable agricultural intensification? *Environmental Research Letters*, *12*(9), 095006.
12 <https://doi.org/10.1088/1748-9326/aa768b>
- 13 Sijapati Basnett, B., Elias, M., Ihalainen, M., & Paez Valencia, A. M. (2017). Gender matters in
14 Forest Landscape Restoration: A framework for design and evaluation. Center for International
15 Forestry Research (CIFOR), Bogor, Indonesia. Retrieved from
16 [https://www.cifor.org/library/6685/gender-matters-in-forest-landscape-restoration-a-framework-
for-design-and-evaluation/](https://www.cifor.org/library/6685/gender-matters-in-forest-landscape-restoration-a-framework-
17 for-design-and-evaluation/)
- 18 Sikder, M. J. U., & Higgins, V. (2017). Remittances and social resilience of migrant households in
19 rural Bangladesh. *Migration and Development*, *6*(2), 253–275.
20 <https://doi.org/10.1080/21632324.2016.1142752>
- 21 Simon, G. L., & Peterson, C. (2018). Disingenuous forests: A historical political ecology of fuelwood
22 collection in South India. *Journal of Historical Geography*.
23 <https://doi.org/10.1016/J.JHG.2018.09.003>
- 24 Simonneaux, V., Cheggour, A., Deschamps, C., Mouillot, F., Cerdan, O., & Le Bissonnais, Y. (2015).
25 Land use and climate change effects on soil erosion in a semi-arid mountainous watershed (High
26 Atlas, Morocco). *Journal of Arid Environments*, *122*, 64–75.
27 <https://doi.org/10.1016/j.jaridenv.2015.06.002>
- 28 Singh, G. (2009). Salinity-related desertification and management strategies: Indian experience. *Land
29 Degradation & Development*, *20*(4), 367–385. <https://doi.org/10.1002/ldr.933>
- 30 Singh, G., Khan, A. U., Kumar, A., Bala, N., Tomar, U. K., G. Singh, ... Tomar, U. K. (2012). Effects
31 of rainwater harvesting and afforestation on soil properties and growth of *Embllica officinalis*
32 while restoring degraded hills in western India. *African Journal of Environmental Science and
33 Technology*, *6*(8), 300–311. <https://doi.org/10.5897/AJEST11.040>
- 34 Singh, K. P., Mohan, D., Sinha, S., & Dalwani, R. (2004). Impact assessment of treated/untreated
35 wastewater toxicants discharged by sewage treatment plants on health, agricultural, and
36 environmental quality in the wastewater disposal area. *Chemosphere*, *55*(2), 227–255.
37 <https://doi.org/10.1016/J.CHEMOSPHERE.2003.10.050>
- 38 Sivakumar, M.V.K., Ndiang'ui, N. (Ed.). (2007). *Climate and Land Degradation (Environmental
39 Science and Engineering)*. Springer.
- 40 Sivakumar, M. V. K. (2007). Interactions between climate and desertification. *Agricultural and
41 Forest Meteorology*, *142*(2–4), 143–155. <https://doi.org/10.1016/j.agrformet.2006.03.025>
- 42 Sloat, L. L., Gerber, J. S., Samberg, L. H., Smith, W. K., Herrero, M., Ferreira, L. G., ... West, P. C.
43 (2018). Increasing importance of precipitation variability on global livestock grazing lands.
44 *Nature Climate Change*, *8*(3), 214–218. <https://doi.org/10.1038/s41558-018-0081-5>
- 45 Smit, G. N., de Klerk, J. N., Schneider, M. B., & van Eck, J. (2015). *Detailed assessment of the
46 biomass resource and potential yield in a selected bush encroached area of Namibia*. Windhoek,
47 Namibia.
- 48 Smith, M. ., & Foran, B. (1992). An approach to assessing the economic risk of different drought

- 1 management tactics on a South Australian pastoral sheep station. *Agricultural Systems*, 39, 83–
2 105.
- 3 Smith, P., Haberl, H., Popp, A., Erb, K., Lauk, C., Harper, R., ... Rose, S. (2013). How much land-
4 based greenhouse gas mitigation can be achieved without compromising food security and
5 environmental goals? *Global Change Biology*, 19(8), 2285–2302.
6 <https://doi.org/10.1111/gcb.12160>
- 7 Smucker, T. A., Wisner, B., Mascarenhas, A., Munishi, P., Wangui, E. E., Sinha, G., ... Lovell, E.
8 (2015). Differentiated livelihoods, local institutions, and the adaptation imperative: Assessing
9 climate change adaptation policy in Tanzania. *Geoforum*, 59, 39–50.
10 <https://doi.org/10.1016/j.geoforum.2014.11.018>
- 11 Snorek, J., Renaud, F. G., & Kloos, J. (2014). Divergent adaptation to climate variability: A case
12 study of pastoral and agricultural societies in Niger. *Global Environmental Change*, 29, 371–
13 386. <https://doi.org/10.1016/J.GLOENVCHA.2014.06.014>
- 14 Solmon, F., Elguindi, N., & Mallet, M. (2012). Radiative and climatic effects of dust over West
15 Africa, as simulated by a regional climate model. *Climate Research*, 52, 97–113.
16 <https://doi.org/10.3354/cr01039>
- 17 Sombrero, A., & De Benito, A. (2010). Carbon accumulation in soil. Ten-year study of conservation
18 tillage and crop rotation in a semi-arid area of Castile-Leon, Spain. *Soil & Tillage Research*,
19 107, 64–70. <https://doi.org/10.1016/j.still.2010.02.009>
- 20 Sommer, S., Zucca, C., Grainger, A., Cherlet, M., Zougmore, R., Sokona, Y., ... Wang, G. (2011).
21 Application of indicator systems for monitoring and assessment of desertification from national
22 to global scales. *L. Degrad. Dev*, 22(2), 184–197. <https://doi.org/10.1002/ldr.1084>
- 23 Sonneveld, B. G. J. S., & Dent, D. L. (2007). How good is GLASOD? *Journal of Environmental*
24 *Management*, 90(1), 274–283. <https://doi.org/10.1016/J.JENVMAN.2007.09.008>
- 25 Souissi, I., Boisson, J. M., Mekki, I., Therond, O., Flichman, G., Wery, J., & Belhouchette, H. (2018).
26 Impact assessment of climate change on farming systems in the South Mediterranean area: a
27 Tunisian case study. *Regional Environmental Change*, 18(3), 637–650.
28 <https://doi.org/10.1007/s10113-017-1130-8>
- 29 Sow, S., Nkonya, E., Meyer, S., & Kato, E. (2016). Cost, Drivers and Action Against Land
30 Degradation in Senegal. In *Economics of Land Degradation and Improvement – A Global*
31 *Assessment for Sustainable Development* (pp. 577–608). Cham: Springer International
32 Publishing. https://doi.org/10.1007/978-3-319-19168-3_19
- 33 Spaliviero, M., De Dapper, M., & Maló, S. (2014). Flood analysis of the Limpopo River basin
34 through past evolution reconstruction and a geomorphological approach. *Natural Hazards and*
35 *Earth System Sciences*, 14(8), 2027–2039. <https://doi.org/10.5194/nhess-14-2027-2014>
- 36 Spaliviero, M., De Dapper, M., Mannaerts, C. M., & Yachan, A. (2011). Participatory approach for
37 integrated basin planning with focus on disaster risk reduction: The Case of the Limpopo River.
38 *Water (Switzerland)*, 3(3), 737–763. <https://doi.org/10.3390/w3030737>
- 39 Spinoni, J., Barbosa, P., De Jager, A., McCormick, N., Naumann, G., Vogt, J. V., ... Mazzeschi, M.
40 (2019). A new global database of meteorological drought events from 1951 to 2016. *Journal of*
41 *Hydrology: Regional Studies*, 22, 100593. <https://doi.org/10.1016/J.EJRH.2019.100593>
- 42 Spinoni, J., Vogt, J., Naumann, G., Carrao, H., & Barbosa, P. (2015). Towards identifying areas at
43 climatological risk of desertification using the Köppen-Geiger classification and FAO aridity
44 index. *International Journal of Climatology*, 35(9), 2210–2222. <https://doi.org/10.1002/joc.4124>
- 45 Sprigg, W. A. (2016). Dust Storms, Human Health and a Global Early Warning System (pp. 59–87).
46 Springer, Cham. https://doi.org/10.1007/978-3-319-30626-1_4
- 47 Squeo, F. A., Aravena, R., Aguirre, E., Pollastri, A., Jorquera, C. B., & Ehleringer, J. R. (2006).
48 Groundwater dynamics in a coastal aquifer in north-central Chile: Implications for groundwater

- 1 recharge in an arid ecosystem. *Journal of Arid Environments*, 67(2), 240–254.
2 <https://doi.org/10.1016/j.jaridenv.2006.02.012>
- 3 Srinivasa Rao, C., Sumantra, K., & Thakur, P. (2015). Carbon Sequestration through Conservation
4 Agriculture in Rainfed Systems. In M. Rajkhowa, D.J., Das, Anup, Ngachan, S.V., Sikka, A.K.
5 and Lyngdoh (Ed.), *Integrated Soil and Water Resource Management for Livelihood and*
6 *Environmental Security* (pp. 56–67). Meghalaya, India: ICAR Research Complex for NEH
7 Region, Umiam.
- 8 Stafford-Smith, M., Griggs, D., Gaffney, O., Ullah, F., Reyers, B., Kanie, N., ... O'Connell, D.
9 (2017). Integration: the key to implementing the Sustainable Development Goals. *Sustainability*
10 *Science*, 12(6), 911–919. <https://doi.org/10.1007/s11625-016-0383-3>
- 11 Stafford Smith, M. (2016). Desertification: Reflections on the Mirage (pp. 539–560). Springer, Berlin,
12 Heidelberg. https://doi.org/10.1007/978-3-642-16014-1_20
- 13 Stahle, D. W., Cleaveland, M. K., Grissino-Mayer, H. D., Griffin, R. D., Fye, F. K., Therrell, M.
14 D., ... Diaz, J. V. (2009). Cool- and Warm-Season Precipitation Reconstructions over Western
15 New Mexico. *Journal of Climate*, 22(13), 3729–3750. <https://doi.org/10.1175/2008JCLI2752.1>
- 16 Stambouli, A., Khiat, Z., Flazi, S., & Y Kitamura. (2012). A review on the renewable energy
17 development in Algeria: Current perspective, energy scenario and sustainability issues.
18 *Renewable and Sustainable Energy Reviews*, 16(7). Retrieved from
19 <http://www.sciencedirect.com/science/article/pii/S1364032112003061>
- 20 State Forestry Administration of China. (2015). *A bulletin of desertification and sandification state of*
21 *China*. Retrieved from [http://hmfz.forestry.gov.cn/uploadfile/main/2015-12/file/2015-12-29-](http://hmfz.forestry.gov.cn/uploadfile/main/2015-12/file/2015-12-29-3264a2babd924d22995dd5ed3602bccb.pdf)
22 [3264a2babd924d22995dd5ed3602bccb.pdf](http://hmfz.forestry.gov.cn/uploadfile/main/2015-12/file/2015-12-29-3264a2babd924d22995dd5ed3602bccb.pdf)
- 23 Staver, A. C., Archibald, S., & Levin, S. A. (2011). The global extent and determinant of savanna and
24 forest as alternative biome states. *Science*, 334(October), 230–233.
- 25 Stavi, I., Shem-Tov, R., Ragolsky, G., Lekach, J., Stavi, I., Shem-Tov, R., ... Lekach, J. (2017).
26 Ancient to Recent-Past Runoff Harvesting Agriculture in Recharge Playas of the Hyper-Arid
27 Southern Israel. *Water*, 9(12), 991. <https://doi.org/10.3390/w9120991>
- 28 Stefanski, R., & Sivakumar, M. V. K. (2009). Impacts of sand and dust storms on agriculture and
29 potential agricultural applications of a SDSWS. *IOP Conference Series: Earth and*
30 *Environmental Science*, 7(1), 012016. <https://doi.org/10.1088/1755-1307/7/1/012016>
- 31 Steger, C., Butt, B., & Hooten, M. B. (2017). Safari Science: assessing the reliability of citizen
32 science data for wildlife surveys. *Journal of Applied Ecology*, 54(6), 2053–2062.
33 <https://doi.org/10.1111/1365-2664.12921>
- 34 Sterk, G., Boardman, J., & Verdoodt, A. (2016). Desertification: History, Causes and Options for Its
35 Control. *Land Degradation & Development*, 27(8), 1783–1787. <https://doi.org/10.1002/ldr.2525>
- 36 Stern, D. I. (2017). The environmental Kuznets curve after 25 years. *Journal of Bioeconomics*, 19(1),
37 7–28. <https://doi.org/10.1007/s10818-017-9243-1>
- 38 Stevens, N., Lehmann, C. E. R., Murphy, B. P., & Durigan, G. (2017). Savanna woody encroachment
39 is widespread across three continents. *Global Change Biology*, 23(1), 235–244.
40 <https://doi.org/10.1111/gcb.13409>
- 41 Stringer, L. C., Dougill, A. J., Mkwambisi, D. D., Dyer, J. C., Kalaba, F. K., & Mngoli, M. (2012).
42 Challenges and opportunities for carbon management in Malawi and Zambia. *Carbon*
43 *Management*, 3(2), 159–173. <https://doi.org/10.4155/cmt.12.14>
- 44 Stringer, L. C., Reed, M. S., Fleskens, L., Thomas, R. J., Le, Q. B., & Lala-Pritchard, T. (2017). A
45 New Dryland Development Paradigm Grounded in Empirical Analysis of Dryland Systems
46 Science. *Land Degradation & Development*, 28(7), 1952–1961. <https://doi.org/10.1002/ldr.2716>
- 47 Stringer, L. C., Twyman, C., & Thomas, D. S. G. (2007). Combating land degradation through

- 1 participatory means: the case of Swaziland. *Ambio*, 36(5), 387–93. Retrieved from
2 <http://www.ncbi.nlm.nih.gov/pubmed/17847803>
- 3 Suleimenov, M., Kaskarbayev, Z., Akshalov, K., & Yushchenko, N. (2014). Conservation Agriculture
4 for Long-Term Soil Productivity (pp. 441–454). Springer, Cham. https://doi.org/10.1007/978-3-319-01017-5_26
- 6 Sultana, F. (2014). Gendering Climate Change: Geographical Insights. *The Professional Geographer*,
7 66(3), 372–381. <https://doi.org/10.1080/00330124.2013.821730>
- 8 Sun, H., Pan, Z., & Liu, X. (2012). Numerical simulation of spatial-temporal distribution of dust
9 aerosol and its direct radiative effects on East Asian climate. *Journal of Geophysical Research:*
10 *Atmospheres*, 117(D13), n/a-n/a. <https://doi.org/10.1029/2011JD017219>
- 11 Swann, A. L. S., Hoffman, F. M., Koven, C. D., & Randerson, J. T. (2016). Plant responses to
12 increasing CO2 reduce estimates of climate impacts on drought severity. *Proceedings of the*
13 *National Academy of Sciences of the United States of America*, 113(36), 10019–24.
14 <https://doi.org/10.1073/pnas.1604581113>
- 15 Swemmer, A. M., Mashele, M., & Ndhlovu, P. D. (2018). Evidence for ecological sustainability of
16 fuelwood harvesting at a rural village in South Africa. *Regional Environmental Change*, 1–11.
17 <https://doi.org/10.1007/s10113-018-1402-y>
- 18 Swift, J. (1988). *Major Issues in Pastoral Development with Special Emphasis on Selected African*
19 *Countries*. Rome, Italy: FAO.
- 20 Swift J. (1996). Desertification: Narratives, winners and losers. In M. Leach and R. Mearns (Ed.), *The*
21 *lie of the land – challenging received wisdom on the African environment*. (pp. 73–90). London:
22 International African Institute. Retrieved from <https://ci.nii.ac.jp/naid/10025956836/>
- 23 Switzman, H., Salem, B., Gad, M., Adeel, Z., & Coulibaly, P. (2018). Conservation planning as an
24 adaptive strategy for climate change and groundwater depletion in Wadi El Natrun, Egypt.
25 *Hydrogeology Journal*, 26(3), 689–703. <https://doi.org/10.1007/s10040-017-1669-y>
- 26 Symeonakis, E., Karathanasis, N., Koukoulas, S., & Panagopoulos, G. (2016). Monitoring Sensitivity
27 to Land Degradation and Desertification with the Environmentally Sensitive Area Index: The
28 Case of Lesvos Island. *Land Degradation & Development*, 27(6), 1562–1573.
29 <https://doi.org/10.1002/ldr.2285>
- 30 Tabieh, M., Al-Karablieh, E., Salman, A., Qtaishat, T., Al-Qudah, H., Majdalawi, M., & Al-Khatib,
31 H. (2015). Economic Analysis of Micro-Catchment Rainwater Harvesting Techniques in
32 Jordan’s Arid Zones. *International Journal of Applied Environmental Sciences*, 10(4), 1205–
33 1225.
- 34 Tadesse, M. A., Shiferaw, B. A., & Erenstein, O. (2015). Weather index insurance for managing
35 drought risk in smallholder agriculture: lessons and policy implications for sub-Saharan Africa.
36 *Agricultural and Food Economics*, 3(1), 26. <https://doi.org/10.1186/s40100-015-0044-3>
- 37 Tal, A. (2016). Rethinking the sustainability of Israel’s irrigation practices in the Drylands. *Water*
38 *Research*, 90, 387–394. <https://doi.org/10.1016/J.WATRES.2015.12.016>
- 39 Tamado, T., Schutz, W., & Milberg, P. (2002). Germination ecology of the weed Parthenium
40 hysterophorus in eastern Ethiopia. *Annals of Applied Biology*, 140(3), 263–270.
41 <https://doi.org/10.1111/j.1744-7348.2002.tb00180.x>
- 42 Tambo, J. A. (2016). Adaptation and resilience to climate change and variability in north-east Ghana.
43 *International Journal of Disaster Risk Reduction*, 17, 85–94.
44 <https://doi.org/10.1016/J.IJDRR.2016.04.005>
- 45 Tambo, J. A., & Wünscher, T. (2015). Identification and prioritization of farmers’ innovations in
46 northern Ghana. *Renewable Agriculture and Food Systems*, 30(06), 537–549.
47 <https://doi.org/10.1017/S1742170514000374>

- 1 Tamou, C., Ripoll-Bosch, R., de Boer, I. J. M. M., & Oosting, S. J. (2018). Pastoralists in a changing
2 environment: The competition for grazing land in and around the W Biosphere Reserve, Benin
3 Republic. *Ambio*, 47(3), 340–354. <https://doi.org/10.1007/s13280-017-0942-6>
- 4 Tan, M., & Li, X. (2015). Does the Green Great Wall effectively decrease dust storm intensity in
5 China? A study based on NOAA NDVI and weather station data. *Land Use Policy*, 43, 42–47.
6 <https://doi.org/10.1016/J.LANDUSEPOL.2014.10.017>
- 7 Tanarhte, M., Hadjinicolaou, P., & Lelieveld, J. (2015). Heat wave characteristics in the eastern
8 Mediterranean and Middle East using extreme value theory. *Climate Research*, 63(2), 99–113.
9 <https://doi.org/10.3354/cr01285>
- 10 Tang, Z., An, H., Zhu, G., & Shangguan, Z. (2018). Beta diversity diminishes in a chronosequence of
11 desertification in a desert steppe. *Land Degradation and Development*, 29(3), 543–550.
12 <https://doi.org/10.1002/ldr.2885>
- 13 Tanveer, M., Anjum, S. A., Hussain, S., Cerdà, A., & Ashraf, U. (2017). Relay cropping as a
14 sustainable approach: problems and opportunities for sustainable crop production.
15 *Environmental Science and Pollution Research*, 24(8), 6973–6988.
16 <https://doi.org/10.1007/s11356-017-8371-4>
- 17 Tao, W. (2014). Aeolian desertification and its control in Northern China. *International Soil and*
18 *Water Conservation Research*, 2(4), 34–41. [https://doi.org/10.1016/S2095-6339\(15\)30056-3](https://doi.org/10.1016/S2095-6339(15)30056-3)
- 19 Tarawneh, Q., & Chowdhury, S. (2018). Trends of Climate Change in Saudi Arabia: Implications on
20 Water Resources. *Climate*, 6(1), 8. <https://doi.org/10.3390/cli6010008>
- 21 Taye, G., Poesen, J., Vanmaercke, M., van Wesemael, B., Martens, L., Teka, D., ... Hallet, V. (2015).
22 Evolution of the effectiveness of stone bunds and trenches in reducing runoff and soil loss in the
23 semi-arid Ethiopian highlands. *Zeitschrift Für Geomorphologie*, 59(4), 477–493.
24 <https://doi.org/10.1127/zfg/2015/0166>
- 25 Taylor, K., Brummer, T., Rew, L. J., Lavin, M., & Maxwell, B. D. (2014). *Bromus tectorum*
26 Response to Fire Varies with Climate Conditions. *Ecosystems*, 17(6), 960–973.
27 <https://doi.org/10.1007/s10021-014-9771-7>
- 28 Taylor, M. J., Aguilar-Støen, M., Castellanos, E., Moran-Taylor, M. J., & Gerkin, K. (2016).
29 International migration, land use change and the environment in Ixcán, Guatemala. *Land Use*
30 *Policy*, 54, 290–301. <https://doi.org/10.1016/J.LANDUSEPOL.2016.02.024>
- 31 Teka, K. (2016). *Parthenium Hysterophorus* (Asteraceae) Expansion, Environmental Impact and
32 Controlling Strategies in Tigray, Northern Ethiopia: A Review. *Journal of Drylands*, 6(1), 434–
33 448. Retrieved from [http://www.mu.edu.et/jd/pdfs/V6N1/Paper 4.pdf](http://www.mu.edu.et/jd/pdfs/V6N1/Paper%204.pdf)
- 34 Tenge, A. J., De Graaff, J., & Hella, J. P. (2004). Social and economic factors affecting the adoption
35 of soil and water conservation in West Usambara highlands, Tanzania. *Land Degradation &*
36 *Development*, 15(2), 99–114. <https://doi.org/10.1002/ldr.606>
- 37 Terink, W., Immerzeel, W. W., & Droogers, P. (2013). Climate change projections of precipitation
38 and reference evapotranspiration for the Middle East and Northern Africa until 2050.
39 *International Journal of Climatology*, 33(14), 3055–3072. <https://doi.org/10.1002/joc.3650>
- 40 Teshome, A., de Graaff, J., & Kessler, A. (2016). Investments in land management in the north-
41 western highlands of Ethiopia: The role of social capital. *Land Use Policy*, 57, 215–228.
42 <https://doi.org/10.1016/J.LANDUSEPOL.2016.05.019>
- 43 Theisen, O. M. (2017). Climate Change and Violence: Insights from Political Science. *Current*
44 *Climate Change Reports*, 3(4), 210–221. <https://doi.org/10.1007/s40641-017-0079-5>
- 45 Thiombiano, L., & Tourino-Soto, I. (2007). Status and Trends in Land Degradation in Africa. In
46 *Climate and Land Degradation* (pp. 39–53). Berlin, Heidelberg: Springer Berlin Heidelberg.
47 https://doi.org/10.1007/978-3-540-72438-4_2

- 1 Thomas, N., & Nigam, S. (2018). Twentieth-Century Climate Change over Africa: Seasonal
2 Hydroclimate Trends and Sahara Desert Expansion. *Journal of Climate*, 31(9), 3349–3370.
3 <https://doi.org/10.1175/JCLI-D-17-0187.1>
- 4 Thomas, R. (2008). Opportunities to reduce the vulnerability of dryland farmers in Central and West
5 Asia and North Africa to climate change. *Agriculture, Ecosystems & Environment*, 126(1–2).
6 Retrieved from <http://www.sciencedirect.com/science/article/pii/S0167880908000212>
- 7 Thompson-Hall, M., Carr, E. R., & Pascual, U. (2016). Enhancing and expanding intersectional
8 research for climate change adaptation in agrarian settings. *Ambio*, 45(S3), 373–382.
9 <https://doi.org/10.1007/s13280-016-0827-0>
- 10 Thornton, P. K., Ericksen, P. J., Herrero, M., & Challinor, A. J. (2014). Climate variability and
11 vulnerability to climate change: a review. *Global Change Biology*, 20(11), 3313–3328.
12 <https://doi.org/10.1111/gcb.12581>
- 13 Thornton, P. K., van de Steeg, J., Notenbaert, A., & Herrero, M. (2009). The impacts of climate
14 change on livestock and livestock systems in developing countries: A review of what we know
15 and what we need to know. *Agricultural Systems*, 101(3), 113–127.
16 <https://doi.org/10.1016/J.AGSY.2009.05.002>
- 17 Tian, F., Brandt, M., Liu, Y. Y., Rasmussen, K., & Fensholt, R. (2017). Mapping gains and losses in
18 woody vegetation across global tropical drylands. *Global Change Biology*, 23(4), 1748–1760.
19 <https://doi.org/10.1111/gcb.13464>
- 20 Tian, F., Brandt, M., Liu, Y. Y., Verger, A., Tagesson, T., Diouf, A. A., ... Fensholt, R. (2016).
21 Remote sensing of vegetation dynamics in drylands: Evaluating vegetation optical depth (VOD)
22 using AVHRR NDVI and in situ green biomass data over West African Sahel. *Remote Sensing
23 of Environment*, 177, 265–276. <https://doi.org/10.1016/J.RSE.2016.02.056>
- 24 Tierney, J. E., Pausata, F. S. R., & deMenocal, P. B. (2017). Rainfall regimes of the Green Sahara.
25 *Science Advances*, 3(1), e1601503. <https://doi.org/10.1126/sciadv.1601503>
- 26 Toderich, K. N., Ismail, S., Juylova, E. A., Rabbimov, A. A., Bekchanov, B. B., Shyuskaya, E. V., ...
27 Radjabov, T. B. (2008). New approaches for biosaline agriculture development, management
28 and conservation of sandy desert ecosystems. In *Biosaline Agriculture and High Salinity
29 Tolerance* (pp. 247–264). Basel: Birkhäuser Basel. [https://doi.org/10.1007/978-3-7643-8554-
30 5_23](https://doi.org/10.1007/978-3-7643-8554-5_23)
- 31 Toderich, K. N., Shyuskaya, E. V., Ismail, S., Gismatullina, L. G., Radjabov, T., Bekchanov, B. B., &
32 Aralova, D. B. (2009). Phytogenic resources of halophytes of Central Asia and their role for
33 rehabilitation of sandy desert degraded rangelands. *Land Degradation & Development*, 20(4),
34 386–396. <https://doi.org/10.1002/ldr.936>
- 35 Torell, L. A. A., Murugan, S., & Ramirez, O. A. A. (2010). Economics of flexible versus conservative
36 stocking strategies to manage climate variability risk. *Rangeland Ecology & Management*, 63,
37 415–425.
- 38 Touré, A., Adamou Didier, T., Rajot, J., Marticorena, B., Bergametti, G., Bouet, C., ... Garba, Z.
39 (2019). *Dynamics of wind erosion and impact of vegetation cover and land use in the Sahel: A
40 case study on sandy dunes in southeastern Niger*. *Catena* (Vol. 177).
41 <https://doi.org/10.1016/j.catena.2019.02.011>
- 42 Traore, S. B., Ali, A., Tinni, S. H., Samake, M., Garba, I., Maigari, I., ... Bouafou, K. G. M. (2014).
43 AGRHYMET: A drought monitoring and capacity building center in the West Africa Region.
44 *Weather and Climate Extremes*, 3, 22–30. <https://doi.org/10.1016/J.WACE.2014.03.008>
- 45 Tsakiris, G. (2017). Facets of Modern Water Resources Management: Prolegomena. *Water Resources
46 Management*, 31(10), 2899–2904. <https://doi.org/10.1007/s11269-017-1742-2>
- 47 Tun, K. K. K., Shrestha, R. P., & Datta, A. (2015). Assessment of land degradation and its impact on
48 crop production in the Dry Zone of Myanmar. *International Journal of Sustainable Development*

- 1 & *World Ecology*, 22(6), 533–544. <https://doi.org/10.1080/13504509.2015.1091046>
- 2 Türkeş, M. (1999). Vulnerability of Turkey to desertification with respect to precipitation and aridity
3 conditions. *Turkish Journal of Engineering and Environmental Sciences*, 23(5), 363–380.
4 Retrieved from <http://journals.tubitak.gov.tr/engineering/abstract.htm?id=3523>
- 5 Türkeş, M. (2003). Spatial and Temporal Variations in Precipitation and Aridity Index Series of
6 Turkey. In *Mediterranean Climate* (pp. 181–213). Berlin, Heidelberg: Springer Berlin
7 Heidelberg. https://doi.org/10.1007/978-3-642-55657-9_11
- 8 Türkeş, M. (2013). *İklim Verileri Kullanılarak Türkiye'nin Çölleşme Haritası Dokümanı*
9 *Hazırlanması Raporu. First*. Ankara: Orman ve Su İşleri Bakanlığı, Çölleşme ve Erozyonla
10 Mücadele Genel Müdürlüğü Yayını. Retrieved from
11 [https://www.researchgate.net/publication/293334692_İklim_Verileri_Kullanılarak_Turkiye'nin_](https://www.researchgate.net/publication/293334692_İklim_Verileri_Kullanılarak_Turkiye'nin_Collesme_Haritasi_Dokumani_Hazirlanmasi_Raporu)
12 [Collesme_Haritasi_Dokumani_Hazirlanmasi_Raporu](https://www.researchgate.net/publication/293334692_İklim_Verileri_Kullanılarak_Turkiye'nin_Collesme_Haritasi_Dokumani_Hazirlanmasi_Raporu)
- 13 Türkeş, M. (2017). Recent spatiotemporal variations of synoptic meteorological sand and dust storm
14 events observed over the Middle East and surrounding regions. In *Proceedings of the 5th*
15 *International Workshop on Sand and Dust Storms (SDS): Dust Sources and their Impacts in the*
16 *Middle East, , 23-25 October 2017* (pp. 45–59). Istanbul.
- 17 Türkeş, M., & Tatlı, H. (2011). Use of the spectral clustering to determine coherent precipitation
18 regions in Turkey for the period 1929–2007. *International Journal of Climatology*, 31(14),
19 2055–2067. <https://doi.org/10.1002/joc.2212>
- 20 Türkeş, M., Turp, M., An, T., Ozturk, N., & Kurnaz, M. L. (2019). Impacts of Climate Change on
21 Precipitation Climatology and Variability in Turkey. In N. B. Harmancioglu & D. Altinbilek
22 (Eds.), *Water Resources of Turkey*. New York: Springe International Publishing.
23 <https://doi.org/DOI 10.1007/978-3-030-11729-0>
- 24 Turner, B. L., Lambin, E. F., & Reenberg, A. (2007). The emergence of land change science for
25 global environmental change and sustainability. *Proceedings of the National Academy of*
26 *Sciences*, 104(52), 20666–20671. <https://doi.org/10.1073/pnas.0704119104>
- 27 Turner, M. D., McPeak, J. G., Gillin, K., Kitchell, E., & Kimambo, N. (2016). Reconciling Flexibility
28 and Tenure Security for Pastoral Resources: the Geography of Transhumance Networks in
29 Eastern Senegal. *Human Ecology*, 44(2), 199–215. <https://doi.org/10.1007/s10745-016-9812-2>
- 30 Twine, W. C., & Holdo, R. M. (2016). Fuelwood sustainability revisited: integrating size structure
31 and resprouting into a spatially realistic fuelshed model. *Journal of Applied Ecology*, 53(6),
32 1766–1776. <https://doi.org/10.1111/1365-2664.12713>
- 33 Uhe, P., Philip, S., Kew, S., Shah, K., Kimutai, J., Mwangi, E., ... Otto, F. (2018). Attributing drivers
34 of the 2016 Kenyan drought. *International Journal of Climatology*, 38, e554–e568.
35 <https://doi.org/10.1002/joc.5389>
- 36 Ukkola, A. M., Pitman, A. J., Kauwe, M. G. De, Abramowitz, G., Herger, N., Evans, J. P., & Decker,
37 M. (2018). Evaluating CMIP5 model agreement for multiple drought metrics. *Journal of*
38 *Hydrometeorology*, JHM-D-17-0099.1. <https://doi.org/10.1175/JHM-D-17-0099.1>
- 39 Ulambayar, T., Fernández-Giménez, M. E., Baival, B., & Batjav, B. (2017). Social Outcomes of
40 Community-based Rangeland Management in Mongolian Steppe Ecosystems. *Conservation*
41 *Letters*, 10(3), 317–327. <https://doi.org/10.1111/conl.12267>
- 42 Umutoni, C., Ayantunde, A., Turner, M., & Sawadogo, G. J. (2016). Community Participation in
43 Decentralized Management of Natural Resources in the Southern Region of Mali. *Environment*
44 *and Natural Resources Research*, 6(2), 1. <https://doi.org/10.5539/enrr.v6n2p1>
- 45 UN-EMG. (2011). *Global Drylands: A UN system-wide response*. United Nations. Retrieved from
46 <https://www.zaragoza.es/contenidos/medioambiente/onu/issue07/1107-eng.pdf>
- 47 UN. (2015). *Outcomes and policy-oriented recommendations from the*. Ankara, Turkey. Retrieved
48 from https://www.unccd.int/sites/default/files/sessions/documents/ICCD_CSTS-3_7/7eng.pdf

- 1 UNCCD. (1994). *Elaboration of an international convention to combat desertification in countries*
2 *experiencing serious drought and/or desertification, particularly in Africa*. Paris.
3 <https://doi.org/Accessed online: http://www.unccd.int/convention/menu.php>
- 4 UNCCD. (2017). *Global Land Outlook* (First). Bonn, Germany: UNCCD. Retrieved from
5 https://www.unccd.int/sites/default/files/documents/2017-09/GLO_Full_Report_low_res.pdf
- 6 UNEP-GEF. (2008). *Desertification, Rangelands and Water Resources Working Group Final*
7 *Thematic Report*. Retrieved from
8 https://postconflict.unep.ch/publications/afg_tech/theme_02/afg_drwr.pdf
- 9 UNEP. (1992). *World Atlas of Desertification*. (N. M. and D. S. G. Thomas, Ed.). UNEP.
- 10 UNEP. (2006). *Global deserts outlook*. United Nations Environment Programme. Retrieved from
11 <http://wedocs.unep.org/handle/20.500.11822/9581>
- 12 UNEP WMO UNCCD. (2016). *Global Assessment of Sand and Dust Storms*. Retrieved from
13 http://library.wmo.int/opac/index.php?lvl=notice_display&id=19638#.WCll6IVOIaE
- 14 Unnevehr, L. (2015). Food safety in developing countries: Moving beyond exports. *Global Food*
15 *Security*, 4, 24–29. <https://doi.org/10.1016/J.GFS.2014.12.001>
- 16 Uno, I., Eguchi, K., Yumimoto, K., Takemura, T., Shimizu, A., Uematsu, M., ... Sugimoto, N.
17 (2009). Asian dust transported one full circuit around the globe. *Nature Geoscience*, 2(8), 557–
18 560. <https://doi.org/10.1038/ngeo583>
- 19 Vachtman, D., Sandler, A., Greenbaum, N., & Herut, B. (2013). Dynamics of suspended sediment
20 delivery to the Eastern Mediterranean continental shelf. *Hydrological Processes*, 27(7), 1105–
21 1116. <https://doi.org/10.1002/hyp.9265>
- 22 Vaidyanathan, A. (2006). Restructuring watershed development programmes. *Economic and Political*
23 *Weekly*, 2984–2987.
- 24 van Andel, T., & Carvalheiro, L. G. (2013). Why urban citizens in developing countries use
25 traditional medicines: the case of suriname. *Evidence-Based Complementary and Alternative*
26 *Medicine : ECAM*, 2013, 687197. <https://doi.org/10.1155/2013/687197>
- 27 Van Auken, O. W. (2009). Causes and consequences of woody plant encroachment into western
28 North American grasslands. *Journal of Environmental Management*, 90(10), 2931–2942.
29 <https://doi.org/10.1016/J.JENVMAN.2009.04.023>
- 30 van der Esch S, B, ten B., E, S., M, B., A, S., A, B., ... van den Berg, M. (2017). *Exploring future*
31 *changes in land use and land condition and the impacts on food, water, climate change and*
32 *biodiversity: Scenarios for the UNCCD Global Land Outlook*. *UNCCD Global Land Outlook*.
- 33 van der Zaag, P., & Savenije, H. (1999). *The management of international waters in EU and SADC*
34 *compared. Physics and Chemistry of the Earth, Part B: Hydrology, Oceans and Atmosphere*
35 (Vol. 24). [https://doi.org/10.1016/S1464-1909\(99\)00048-9](https://doi.org/10.1016/S1464-1909(99)00048-9)
- 36 van Loo, M., Duser, B., Verstraeten, G., Renssen, H., Notebaert, B., D'Haen, K., & Bakker, J. (2017).
37 Human induced soil erosion and the implications on crop yield in a small mountainous
38 Mediterranean catchment (SW-Turkey). *CATENA*, 149, 491–504.
39 <https://doi.org/10.1016/J.CATENA.2016.08.023>
- 40 van Oijen, M., Bellocchi, G., & Höglind, M. (2018). Effects of Climate Change on Grassland
41 Biodiversity and Productivity: The Need for a Diversity of Models. *Agronomy*, 8(2), 14.
42 <https://doi.org/10.3390/agronomy8020014>
- 43 van Rijn, F., Bulte, E., & Adegunle, A. (2012). Social capital and agricultural innovation in Sub-
44 Saharan Africa. *Agricultural Systems*, 108, 112–122.
45 <https://doi.org/10.1016/J.AGSY.2011.12.003>
- 46 van Vuuren, D. P., Sala, O. E., & Pereira, H. M. (2006). The future of vascular plant diversity under
47 four global scenarios. *Ecology and Society*, 11(2), 25. <https://doi.org/25>

- 1 van Wilgen, B. W., Forsyth, G. G., de Klerk, H., Das, S., Khuluse, S., & Schmitz, P. (2010). Fire
2 management in Mediterranean-climate shrublands: a case study from the Cape fynbos, South
3 Africa. *Journal of Applied Ecology*, 47(3), 631–638. Retrieved from
4 <http://dx.doi.org/10.1111/j.1365-2664.2010.01800.x>
- 5 van Zanten, B. T., Verburg, P. H., Espinosa, M., Gomez-y-Paloma, S., Galimberti, G., Kantelhardt,
6 J., ... Viaggi, D. (2014). European agricultural landscapes, common agricultural policy and
7 ecosystem services: a review. *Agronomy for Sustainable Development*, 34(2), 309–325.
8 <https://doi.org/10.1007/s13593-013-0183-4>
- 9 Varga, G., Újvári, G., & Kovács, J. (2014). Spatiotemporal patterns of Saharan dust outbreaks in the
10 Mediterranean Basin. *Aeolian Research*, 15, 151–160.
11 <https://doi.org/10.1016/j.aeolia.2014.06.005>
- 12 Vargas, R., Collins, S. L., Thomey, M. L., Johnson, J. E., Brown, R. F., Natvig, D. O., & Friggens, M.
13 T. (2012). Precipitation variability and fire influence the temporal dynamics of soil CO₂ efflux
14 in an arid grassland. *Global Change Biology*, 18(4), 1401–1411. <https://doi.org/10.1111/j.1365-2486.2011.02628.x>
- 16 Varghese, N., & Singh, N. P. (2016). Linkages between land use changes, desertification and human
17 development in the Thar Desert Region of India. *Land Use Policy*, 51, 18–25.
18 <https://doi.org/10.1016/J.LANDUSEPOL.2015.11.001>
- 19 Vega, J. A., Fernández, C., & Fonturbel, T. (2005). Throughfall, runoff and soil erosion after
20 prescribed burning in gorse shrubland in Galicia (NW Spain). *Land Degradation and
21 Development*, 16(1), 37–51. <https://doi.org/10.1002/ldr.643>
- 22 Vente, J. de, Reed, M. S., Stringer, L. C., Valente, S., & Newig, J. (2016). How does the context and
23 design of participatory decision making processes affect their outcomes? Evidence from
24 sustainable land management in global drylands. *Ecology and Society*. Resilience Alliance Inc.
25 <https://doi.org/10.2307/26270377>
- 26 Verbesselt, J., H. R., Zeileis, A., & Culvenor, D. (2010). Phenological change detection while
27 accounting for abrupt and gradual trends in satellite image time series. *Remote Sensing of
28 Environment*, 114(12), 2970–2980. <https://doi.org/10.1016/J.RSE.2010.08.003>
- 29 Verbesselt, J., Hyndman, R., Newnham, G., & Culvenor, D. (2010). Detecting trend and seasonal
30 changes in satellite image time series. *Remote Sensing of Environment*, 114(1), 106–115.
31 <https://doi.org/10.1016/J.RSE.2009.08.014>
- 32 Verbist, K., Amani, A., Mishra, A., & Jiménez, B. (2016). Strengthening drought risk management
33 and policy: UNESCO International Hydrological Programme's case studies from Africa and
34 Latin America and the Caribbean. *Water Policy*, 18. <https://doi.org/10.2166/wp.2016.223>
- 35 Verhoeven, H. (2011). Climate Change, Conflict and Development in Sudan: Global Neo-Malthusian
36 Narratives and Local Power Struggles. *Development and Change*, 42(3), 679–707.
37 <https://doi.org/10.1111/j.1467-7660.2011.01707.x>
- 38 Verner, D., Tréguer D, J. R., Jens Christensen, R. M. C. E., Konishi, Y., & Belghazi, S. (2018).
39 *Climate Variability, Drought, and Drought Management in Morocco's Agricultural Sector*.
40 Washington D.C. Retrieved from
41 [http://documents.worldbank.org/curated/en/353801538414553978/Climate-Variability-Drought-
42 and-Drought-Management-in-Moroccos-Agricultural-Sector](http://documents.worldbank.org/curated/en/353801538414553978/Climate-Variability-Drought-and-Drought-Management-in-Moroccos-Agricultural-Sector)
- 43 Verón, S. R., Blanco, L. J., Texeira, M. A., Irisarri, J. G. N., & Puelo, J. M. (2017). Desertification
44 and ecosystem services supply: The case of the Arid Chaco of South America. *Journal of Arid
45 Environments*, 159, 66–74. <https://doi.org/10.1016/J.JARIDENV.2017.11.001>
- 46 Verstraete, M. M., Hutchinson, C. F., Grainger, A., Stafford Smith, M., Scholes, R. J., Reynolds, J.
47 F., ... Mbow, C. (2011). Towards a global drylands observing system: Observational
48 requirements and institutional solutions. *Land Degradation and Development*, 22(2), 198–213.
49 <https://doi.org/10.1002/ldr.1046>

- 1 Vetter, M., Wirth, C., Bottcher, H., Churkina, G., Schulze, E.-D., Wutzler, T., & Weber, G. (2005).
2 Partitioning direct and indirect human-induced effects on carbon sequestration of managed
3 coniferous forests using model simulations and forest inventories. *Global Change Biology*,
4 *11*(5), 810–827. <https://doi.org/10.1111/j.1365-2486.2005.00932.x>
- 5 Vetter, S. (2005). Rangelands at equilibrium and non-equilibrium: recent developments in the debate.
6 *Journal of Arid Environments*, *62*(2), 321–341.
7 <https://doi.org/10.1016/J.JARIDENV.2004.11.015>
- 8 Vicente-Serrano, S. M., Beguería, S., Gimeno, L., Eklundh, L., Giuliani, G., Weston, D., ... Pegram,
9 G. G. S. (2012). Challenges for drought mitigation in Africa: The potential use of geospatial data
10 and drought information systems. *Applied Geography*, *34*, 471–486.
11 <https://doi.org/10.1016/J.APGEOG.2012.02.001>
- 12 Vidadili, N., Suleymanov, E., Bulut, C., & Mahmudlu, C. (2017). Transition to renewable energy and
13 sustainable energy development in Azerbaijan. *Renewable and Sustainable Energy Reviews*, *80*,
14 1153–1161. <https://doi.org/10.1016/J.RSER.2017.05.168>
- 15 Vieira, R. M. S. P., Tomasella, J., Alvalá, R. C. S., Sestini, M. F., Affonso, A. G., Rodriguez, D.
16 A., ... Santana, M. O. (2015). Identifying areas susceptible to desertification in the Brazilian
17 northeast. *Solid Earth*, *6*(1), 347–360. <https://doi.org/10.5194/se-6-347-2015>
- 18 Villamor, G. B., & van Noordwijk, M. (2016). Gender specific land-use decisions and implications
19 for ecosystem services in semi-matrilineal Sumatra. *Global Environmental Change*, *39*, 69–80.
20 <https://doi.org/10.1016/J.GLOENVCHA.2016.04.007>
- 21 Villamor, G. B., van Noordwijk, M., Djanibekov, U., Chiong-Javier, M. E., & Catacutan, D. (2014).
22 Gender differences in land-use decisions: shaping multifunctional landscapes? *Current Opinion*
23 *in Environmental Sustainability*, *6*, 128–133. <https://doi.org/10.1016/J.COSUST.2013.11.015>
- 24 Visser, W. P. (2018). A perfect storm: The ramifications of Cape Town's drought crisis. *The Journal*
25 *for Transdisciplinary Research in Southern Africa*, *14*(1), 10.
26 <https://doi.org/10.4102/td.v14i1.567>
- 27 Vogt, J. V., Safriel, U., Von Maltitz, G., Sokona, Y., Zougmore, R., Bastin, G., & Hill, J. (2011).
28 Monitoring and assessment of land degradation and desertification: Towards new conceptual
29 and integrated approaches. *Land Degradation & Development*, *22*(2), 150–165.
30 <https://doi.org/10.1002/ldr.1075>
- 31 Vohland, K., & Barry, B. (2009). A review of in situ rainwater harvesting (RWH) practices modifying
32 landscape functions in African drylands. *Agriculture, Ecosystems and Environment*, *131*(3–4),
33 119–127. <https://doi.org/10.1016/j.agee.2009.01.010>
- 34 Von Braun, J., & Gatzweiler, F. W. (2014). *Marginality: addressing the nexus of poverty, exclusion*
35 *and ecology*.
- 36 von Uexkull, N., Croicu, M., Fjelde, H., & Buhaug, H. (2016). Civil conflict sensitivity to growing-
37 season drought. *Proceedings of the National Academy of Sciences of the United States of*
38 *America*, *113*(44), 12391–12396. <https://doi.org/10.1073/pnas.1607542113>
- 39 von Wehrden, H., Hanspach, J., Kaczensky, P., Fischer, J., & Wesche, K. (2012). Global assessment
40 of the non-equilibrium concept in rangelands. *Ecological Applications*, *22*(2), 393–399.
41 <https://doi.org/10.1890/11-0802.1>
- 42 Vu, Q. M., Le, Q. B., Frossard, E., & Vlek, P. L. G. (2014). Socio-economic and biophysical
43 determinants of land degradation in Vietnam: An integrated causal analysis at the national level.
44 *Land Use Policy*, *36*, 605–617. <https://doi.org/10.1016/J.LANDUSEPOL.2013.10.012>
- 45 Waha, K., Krummenauer, L., Adams, S., Aich, V., Baarsch, F., Coumou, D., ... Schleussner, C.-F.
46 (2017). Climate change impacts in the Middle East and Northern Africa (MENA) region and
47 their implications for vulnerable population groups. *Regional Environmental Change*, *17*(6),
48 1623–1638. <https://doi.org/10.1007/s10113-017-1144-2>

- 1 Waha, K., Müller, C., Bondeau, A., Dietrich, J. P., Kurukulasuriya, P., Heinke, J., & Lotze-Campen,
2 H. (2013). Adaptation to climate change through the choice of cropping system and sowing date
3 in sub-Saharan Africa. *Global Environmental Change*, 23(1), 130–143.
4 <https://doi.org/10.1016/J.GLOENVCHA.2012.11.001>
- 5 Walther, B. A. (2016). A review of recent ecological changes in the Sahel, with particular reference to
6 land use change, plants, birds and mammals. *African Journal of Ecology*, 54(3), 268–280.
7 <https://doi.org/10.1111/aje.12350>
- 8 Walther, G.-R., Post, E., Convey, P., Menzel, A., Parmesan, C., Beebee, T. J. C., ... Bairlein, F.
9 (2002). Ecological responses to recent climate change. *Nature*, 416(6879), 389–395.
10 <https://doi.org/10.1038/416389a>
- 11 Walther, G. R. (2010). Community and ecosystem responses to recent climate change. *Philosophical
12 Transactions of the Royal Society B: Biological Sciences*, 365(1549), 2019–2024.
13 <https://doi.org/10.1098/rstb.2010.0021>
- 14 Wang, F., Pan, X., Wang, D., Shen, C., & Lu, Q. (2013). Combating desertification in China: Past,
15 present and future. *Land Use Policy*, 31, 311–313.
16 <https://doi.org/10.1016/j.landusepol.2012.07.010>
- 17 Wang, G. (2005). Agricultural drought in a future climate: Results from 15 global climate models
18 participating in the IPCC 4th assessment. *Climate Dynamics*, 25(7–8), 739–753.
19 <https://doi.org/10.1007/s00382-005-0057-9>
- 20 Wang, H., Pan, Y., Chen, Y., & Ye, Z. (2017). Linear trend and abrupt changes of climate indices in
21 the arid region of northwestern China. *Atmospheric Research*, 196, 108–118. Retrieved from
22 <https://www.sciencedirect.com/science/article/pii/S016980951630552X>
- 23 Wang, K., & Dickinson, R. E. (2013). Contribution of solar radiation to decadal temperature
24 variability over land. *Proceedings of the National Academy of Sciences of the United States of
25 America*, 110(37), 14877–82. <https://doi.org/10.1073/pnas.1311433110>
- 26 Wang, X., Hua, T., Lang, L., & Ma, W. (2017). Spatial differences of aeolian desertification
27 responses to climate in arid Asia. *Global and Planetary Change*, 148, 22–28.
28 <https://doi.org/10.1016/j.gloplacha.2016.11.008>
- 29 Wang, X. M., Zhang, C. X., Hasi, E., & Dong, Z. B. (2010). Has the Three Norths Forest Shelterbelt
30 Program solved the desertification and dust storm problems in arid and semiarid China? *Journal
31 of Arid Environments*, 74(1), 13–22. <https://doi.org/10.1016/J.JARIDENV.2009.08.001>
- 32 Wang, X., Yamauchi, F., Huang, J., & Rozelle, S. (2018). What constrains mechanization in Chinese
33 agriculture? Role of farm size and fragmentation. *China Economic Review*.
34 <https://doi.org/10.1016/J.CHIECO.2018.09.002>
- 35 Wang, X., Yamauchi, F., Otsuka, K., & Huang, J. (2014). *Wage Growth, Landholding, and
36 Mechanization in Chinese Agriculture*. The World Bank. [https://doi.org/10.1596/1813-9450-
37 7138](https://doi.org/10.1596/1813-9450-7138)
- 38 Wang, X., Yang, Y., Dong, Z., & Zhang, C. (2009). Responses of dune activity and desertification in
39 China to global warming in the twenty-first century. *Global and Planetary Change*, 67(3–4),
40 167–185. <https://doi.org/10.1016/j.gloplacha.2009.02.004>
- 41 Wang, Y., Xu, X., Lei, J., Li, S., Zhou, Z., Chang, Q., ... Xu, B. (2008). The dynamics variation of
42 soil moisture of shelterbelts along the Tarim Desert Highway. *Science Bulletin*, 53(S2), 102–
43 108. <https://doi.org/10.1007/s11434-008-6011-6>
- 44 Wang, Y., Zhang, J., Tong, S., & Guo, E. (2017). Monitoring the trends of aeolian desertified lands
45 based on time-series remote sensing data in the Horqin Sandy Land, China. *Catena*, 157(May),
46 286–298. <https://doi.org/10.1016/j.catena.2017.05.030>
- 47 Wang, Z., Zhou, B., Pei, L., Zhang, J., He, X., & Lin, H. (2018). Controlling threshold in soil salinity
48 when planting spring wheat and sequential cropping silage corn in Northern Xinjiang using drip

- 1 irrigation. *International Journal of Agricultural and Biological Engineering*, 11(2), 108–114.
2 <https://doi.org/10.25165/IJABE.V11I2.3621>
- 3 Wario, H. T., Roba, H. G., & Kaufmann, B. (2016). Responding to mobility constraints: Recent shifts
4 in resource use practices and herding strategies in the Borana pastoral system, southern Ethiopia.
5 *Journal of Arid Environments*, 127, 222–234. <https://doi.org/10.1016/J.JARIDENV.2015.12.005>
- 6 Waters-Bayer, A., van Veldhuizen, L., Wongtschowski, M., & Wettasinha, C. (2009). Recognizing
7 and enhancing processes of local innovation. In P. Sanginga, A. Waters-Bayer, S. Kaaria, J.
8 Njuki, & C. Wettasinha (Eds.), *Innovation Africa: Enriching Farmers' Livelihoods*. London,
9 UK: Earthscan.
- 10 Way, S.-A. (2016). Examining the Links between Poverty and Land Degradation: From Blaming the
11 Poor toward Recognising the Rights of the Poor, 47–62.
12 <https://doi.org/10.4324/9781315253916-13>
- 13 Webb, N. P., Marshall, N. A., Stringer, L. C., Reed, M. S., Chappell, A., & Herrick, J. E. (2017).
14 Land degradation and climate change: building climate resilience in agriculture. *Frontiers in*
15 *Ecology and the Environment*, 15(8), 450–459. <https://doi.org/10.1002/fee.1530>
- 16 Webb, N. P., Stokes, C. J., & Marshall, N. A. (2013). Integrating biophysical and socio-economic
17 evaluations to improve the efficacy of adaptation assessments for agriculture. *Global*
18 *Environmental Change*, 23(5), 1164–1177.
19 <https://doi.org/https://doi.org/10.1016/j.gloenvcha.2013.04.007>
- 20 Weber, B., Bowker, M., Zhang, Y., & Belnap, J. (2016). Natural Recovery of Biological Soil Crusts
21 After Disturbance. In B. Weber, B. Büdel, & J. Belnap (Eds.), *Biological Soil Crusts: An*
22 *Organizing Principle in Drylands* (pp. 479–498). New York: Springer International Publishing.
23 https://doi.org/10.1007/978-3-319-30214-0_23
- 24 Weber, C. F., Lockhart, J. S., Charaska, E., Aho, K., & Lohse, K. A. (2014). Bacterial composition of
25 soils in ponderosa pine and mixed conifer forests exposed to different wildfire burn severity.
26 *Soil Biology and Biochemistry*, 69, 242–250. <https://doi.org/10.1016/j.soilbio.2013.11.010>
- 27 Weinzierl, T., Wehberg, J., Böhner, J., & Conrad, O. (2016). Spatial Assessment of Land Degradation
28 Risk for the Okavango River Catchment, Southern Africa. *Land Degradation & Development*,
29 27(2), 281–294. <https://doi.org/10.1002/ldr.2426>
- 30 Weltz, M., & Spaeth, K. (2012). Estimating Effects of Targeted Conservation on Nonfederal
31 Rangelands. *Rangelands*, 34(4), 35–40. [https://doi.org/https://doi.org/10.2111/RANGELANDS-](https://doi.org/https://doi.org/10.2111/RANGELANDS-D-12-00028.1)
32 [D-12-00028.1](https://doi.org/10.2111/RANGELANDS-D-12-00028.1)
- 33 Wessels, K. J., Prince, S. D., Malherbe, J., Small, J., Frost, P. E., & VanZyl, D. (2007). Can human-
34 induced land degradation be distinguished from the effects of rainfall variability? A case study
35 in South Africa. *Journal of Arid Environments*, 68(2), 271–297.
36 <https://doi.org/10.1016/J.JARIDENV.2006.05.015>
- 37 Wessels, K. J., van den Bergh, F., & Scholes, R. J. (2012). Limits to detectability of land degradation
38 by trend analysis of vegetation index data. *Remote Sensing of Environment*, 125, 10–22.
39 <https://doi.org/10.1016/J.RSE.2012.06.022>
- 40 White, R., & Nackoney. (2003). *Drylands, people, and ecosystem goods and services: a web-based*
41 *geospatial analysis*. Retrieved from <http://pdf.wri.org/drylands.pdf>
- 42 Whitford, W. G. (1997). Desertification and animal biodiversity in the desert grasslands of North
43 America. *Journal of Arid Environments*, 37(4), 709–720. <https://doi.org/10.1006/jare.1997.0313>
- 44 Whitford, W. G. (Walter G. . (2002). *Ecology of desert systems*. San Diego: Academic Press.
45 Retrieved from <http://www.worldcat.org/title/ecology-of-desert-systems/oclc/49238045>
- 46 Wichelns, D., & Qadir, M. (2015). Achieving sustainable irrigation requires effective management of
47 salts, soil salinity, and shallow groundwater. *Agricultural Water Management*, 157, 31–38.
48 <https://doi.org/10.1016/J.AGWAT.2014.08.016>

- 1 Wicke, B., Smeets, E. M. W., Akanda, R., Stille, L., Singh, R. K., Awan, A. R., ... Faaij, A. P. C.
2 (2013). Biomass production in agroforestry and forestry systems on salt-affected soils in South
3 Asia: Exploration of the GHG balance and economic performance of three case studies. *Journal*
4 *of Environmental Management*, 127, 324–334.
5 <https://doi.org/10.1016/J.JENVMAN.2013.05.060>
- 6 Wiesmeier, M., Munro, S., Barthold, F., Steffens, M., Schad, P., & Kögel-Knabner, I. (2015). Carbon
7 storage capacity of semi-arid grassland soils and sequestration potentials in northern China.
8 *Global Change Biology*, 21(10), 3836–3845. <https://doi.org/10.1111/gcb.12957>
- 9 Wildemeersch, J. C. J., Timmerman, E., Mazijn, B., Sabiou, M., Ibro, G., Garba, M., & Cornelis, W.
10 (2015). Assessing the Constraints to Adopt Water and Soil Conservation Techniques in
11 Tillaberi, Niger. *Land Degradation & Development*, 26(5), 491–501.
12 <https://doi.org/10.1002/ldr.2252>
- 13 Wilhelm, W. W., & Wortmann, C. S. (2004). Tillage and Rotation Interactions for Corn and Soybean
14 Grain Yield as Affected by Precipitation and Air Temperature Joint contribution the USDA-
15 ARS and the Univ. of Nebraska, Agric. Res. Div., Lincoln, NE 68583, Journal Series no. 13950.
16 Product names are. *Agronomy Journal*, 96, 425–432. <https://doi.org/10.2134/agronj2004.4250>
- 17 Wilhite, D. A. (2016). Drought-Management Policies and Preparedness Plans: Changing the Paradigm
18 from Crisis to Risk Management. *Land Restoration*, 443–462. <https://doi.org/10.1016/B978-0-12-801231-4.00007-0>
- 19
- 20 Wilhite, D. A., Sivakumar, M. V. K., & Pulwarty, R. (2014). Managing drought risk in a changing
21 climate: The role of national drought policy. *Weather and Climate Extremes*, 3, 4–13.
22 <https://doi.org/10.1016/J.WACE.2014.01.002>
- 23 Wilhite, D., & Pulwarty, R. S. (2017). *Drought and Water Crises, Integrating Science, Management,*
24 *and Policy* (Second). Boca Raton, Florida: CRC Press. <https://doi.org/10.1201/b22009>
- 25 Williams, A. P., Allen, C. D., Millar, C. I., Swetnam, T. W., Michaelsen, J., Still, C. J., & Leavitt, S.
26 W. (2010). Forest responses to increasing aridity and warmth in the southwestern United States.
27 *Proceedings of the National Academy of Sciences*, 107(50), 21289–21294.
28 <https://doi.org/10.1073/pnas.0914211107>
- 29 Willy, D. K., & Holm-Müller, K. (2013). Social influence and collective action effects on farm level
30 soil conservation effort in rural Kenya. *Ecological Economics*, 90, 94–103.
31 <https://doi.org/10.1016/J.ECOLECON.2013.03.008>
- 32 Wilting, H. C., Schipper, A. M., Bakkenes, M., Meijer, J. R., & Huijbregts, M. A. J. (2017).
33 Quantifying Biodiversity Losses Due to Human Consumption: A Global-Scale Footprint
34 Analysis. *Environmental Science & Technology*, 51(6), 3298–3306.
35 <https://doi.org/10.1021/acs.est.6b05296>
- 36 Winsemius, H. C., Jongman, B., Veldkamp, T. I. E., Hallegatte, S., Bangalore, M., & Ward, P. J.
37 (2018). Disaster risk, climate change, and poverty: assessing the global exposure of poor people
38 to floods and droughts. *Environment and Development Economics*, 23(03), 328–348.
39 <https://doi.org/10.1017/S1355770X17000444>
- 40 Wischnath, G., & Buhaug, H. (2014). Rice or riots: On food production and conflict severity across
41 India. *Political Geography*, 43, 6–15. <https://doi.org/10.1016/J.POLGEO.2014.07.004>
- 42 Womack, A. M., Bohannon, B. J. M., & Green, J. L. (2010). Biodiversity and biogeography of the
43 atmosphere. *Philosophical Transactions of the Royal Society B: Biological Sciences*, 365(1558),
44 3645–3653. <https://doi.org/10.1098/rstb.2010.0283>
- 45 Woo, A. C., Brar, M. S., Chan, Y., Lau, M. C. Y., Leung, F. C. C., Scott, J. A., ... Pointing, S. B.
46 (2013). Temporal variation in airborne microbial populations and microbially-derived allergens
47 in a tropical urban landscape. *Atmospheric Environment*, 74, 291–300.
48 <https://doi.org/10.1016/j.atmosenv.2013.03.047>

- 1 Woodhouse, P., Veldwisch, G. J., Venot, J.-P., Brockington, D., Komakech, H., & Manjichi, Â.
2 (2017). African farmer-led irrigation development: re-framing agricultural policy and
3 investment? *The Journal of Peasant Studies*, 44(1), 213–233.
4 <https://doi.org/10.1080/03066150.2016.1219719>
- 5 Woolcock, M., & Narayan, D. (2000). Social capital: implications for development theory, research,
6 and policy. *The World Bank Research Observer*, 15(2 (August 2000)), 25–249. Retrieved from
7 [http://documents.worldbank.org/curated/en/961231468336675195/Social-capital-implications-](http://documents.worldbank.org/curated/en/961231468336675195/Social-capital-implications-for-development-theory-research-and-policy)
8 [for-development-theory-research-and-policy](http://documents.worldbank.org/curated/en/961231468336675195/Social-capital-implications-for-development-theory-research-and-policy)
- 9 Woollen, E., Ryan, C. M., Grundy, I., Baumert, S., Vollmer, F., Fernando, J., ... Lisboa, S. N. (2016).
10 Supplementary materials for Charcoal production in the Mopane woodlands of Mozambique:
11 what are the tradeoffs with other ecosystem services? *Philosophical Transactions of the Royal*
12 *Society B-Biological Sciences*, 371, 1–14. <https://doi.org/10.1098/rstb.2015.0315>
- 13 Woomer, P. L., Touré, A., & Sall, M. (2004). Carbon stocks in Senegal's Sahel Transition Zone.
14 *Journal of Arid Environments*, 59, 499–510.
- 15 World Bank. (2009, November 6). World Development Report 2010: Development and Climate
16 Change. <https://doi.org/10.1596/978-0-8213-7987-5>
- 17 World Bank. (2018). *Groundswell: preparing for internal climate migration*. Washington: World
18 Bank. Retrieved from [http://documents.worldbank.org/curated/en/846391522306665751/Main-](http://documents.worldbank.org/curated/en/846391522306665751/Main-report)
19 [report](http://documents.worldbank.org/curated/en/846391522306665751/Main-report)
- 20 Wossen, T., Berger, T., & Di Falco, S. (2015). Social capital, risk preference and adoption of
21 improved farm land management practices in Ethiopia. *Agricultural Economics*, 46(1), 81–97.
22 <https://doi.org/10.1111/agec.12142>
- 23 Wu, J., Zhou, L., Mo, X., Zhou, H., Zhang, J., & Jia, R. (2015). Drought monitoring and analysis in
24 China based on the Integrated Surface Drought Index (ISDI). *International Journal of Applied*
25 *Earth Observation and Geoinformation*, 41, 23–33. <https://doi.org/10.1016/J.JAG.2015.04.006>
- 26 Wu, W. (2014). The Generalized Difference Vegetation Index (GDVI) for Dryland Characterization.
27 *Remote Sensing*, 6(2), 1211–1233. <https://doi.org/10.3390/rs6021211>
- 28 Xi, X., & Sokolik, I. N. (2015). Seasonal dynamics of threshold friction velocity and dust emission in
29 Central Asia. *Journal of Geophysical Research: Atmospheres*, 120(4), 1536–1564.
30 <https://doi.org/10.1002/2014JD022471>
- 31 Xie, Y., Gong, J., Sun, P., & Gou, X. (2014). Oasis dynamics change and its influence on landscape
32 pattern on Jinta oasis in arid China from 1963a to 2010a: Integration of multi-source satellite
33 images. *International Journal of Applied Earth Observation and Geoinformation*, 33, 181–191.
34 Retrieved from <https://linkinghub.elsevier.com/retrieve/pii/S0303243414001287>
- 35 Xu, D. Y. Y., Kang, X. W. W., Zhuang, D. F. F., & Pan, J. J. J. (2010). Multi-scale quantitative
36 assessment of the relative roles of climate change and human activities in desertification - A
37 case study of the Ordos Plateau, China. *Journal of Arid Environments*, 74(4), 498–507.
38 <https://doi.org/10.1016/j.jaridenv.2009.09.030>
- 39 Xu, D., You, X., & Xia, C. (2019). Assessing the spatial-temporal pattern and evolution of areas
40 sensitive to land desertification in North China. *Ecological Indicators*, 97, 150–158.
41 <https://doi.org/10.1016/J.ECOLIND.2018.10.005>
- 42 Xue, Z., Qin, Z., Cheng, F., Ding, G., & Li, H. (2017). Quantitative assessment of aeolian
43 desertification dynamics– A case study in north Shanxi of China (1975 to 2015). *Scientific*
44 *Reports*, 7(1), 10460. <https://doi.org/10.1038/s41598-017-11073-8>
- 45 Yahiya, A. B. (2012). Environmental degradation and its impact on tourism in Jazan , KSA using
46 Remote Sensing and GIS, 3(1), 421–432. <https://doi.org/10.6088/ijes.2012030131041>
- 47 Yaish, M. W., & Kumar, P. P. (2015). Salt tolerance research in date palm tree (*Phoenix dactylifera*
48 L.), past, present, and future perspectives. *Frontiers in Plant Science*, 6, 348.

- 1 <https://doi.org/10.3389/fpls.2015.00348>
- 2 Yan, X., & Cai, Y. L. (2015). Multi-Scale Anthropogenic Driving Forces of Karst Rocky
3 Desertification in Southwest China. *Land Degradation & Development*, 26(2), 193–200.
4 <https://doi.org/10.1002/ldr.2209>
- 5 Yang, D., Kanae, S., Oki, T., Koike, T., & Musiak, K. (2003). Global potential soil erosion with
6 reference to land use and climate changes. *Hydrological Processes*, 17(14), 2913–2928.
7 <https://doi.org/10.1002/hyp.1441>
- 8 Yang, F., & Lu, C. (2015). Spatiotemporal variation and trends in rainfall erosivity in China's dryland
9 region during 1961-2012. *Catena*, 133, 362–372. <https://doi.org/10.1016/j.catena.2015.06.005>
- 10 Yang, H., Li, X., Wang, Z., Jia, R., Liu, L., Chen, Y., ... Li, G. (2014). Carbon sequestration capacity
11 of shifting sand dune after establishing new vegetation in the Tengger Desert, northern China.
12 *Science of The Total Environment*, 478, 1–11.
13 <https://doi.org/10.1016/J.SCITOTENV.2014.01.063>
- 14 Yang, L., Wu, J., & Shen, P. (2013). Roles of science in institutional changes: the case of
15 desertification control in China. *Environmental Science & Policy*, 27, 32–54. Retrieved from
16 <http://www.sciencedirect.com/science/article/pii/S1462901112001888>
- 17 Yang, Y., Wang, Z., Li, J., Gang, C., Zhang, Y., Zhang, Y., ... Qi, J. (2016). Comparative assessment
18 of grassland degradation dynamics in response to climate variation and human activities in
19 China, Mongolia, Pakistan and Uzbekistan from 2000 to 2013. *Journal of Arid Environments*,
20 135, 164–172. <https://doi.org/10.1016/J.JARIDENV.2016.09.004>
- 21 Yanoff, S., & Muldavin, E. (2008). Grassland–shrubland transformation and grazing: A century-scale
22 view of a northern Chihuahuan Desert grassland. *Journal of Arid Environments*, 72(9), 1594–
23 1605. <https://doi.org/10.1016/J.JARIDENV.2008.03.012>
- 24 Yaro, J. A., Teye, J., & Bawakyillenuo, S. (2015). Local institutions and adaptive capacity to climate
25 change/variability in the northern savannah of Ghana. *Climate and Development*, 7(3), 235–245.
26 <https://doi.org/10.1080/17565529.2014.951018>
- 27 Ye, X., Yu, X., Yu, C., Tayibzhaer, A., Xu, F., Skidmore, A. K., & Wang, T. (2018). Impacts of
28 future climate and land cover changes on threatened mammals in the semi-arid Chinese Altai
29 Mountains. *Science of The Total Environment*, 612, 775–787.
30 <https://doi.org/10.1016/J.SCITOTENV.2017.08.191>
- 31 Yi, C., Wei, S., & Hendrey, G. (2014). Warming climate extends dryness-controlled areas of
32 terrestrial carbon sequestration. *Scientific Reports*, 4, 1–6. <https://doi.org/10.1038/srep05472>
- 33 Yin, F., Deng, X., Jin, Q., Yuan, Y., & Zhao, C. (2014). The impacts of climate change and human
34 activities on grassland productivity in Qinghai Province, China. *Frontiers of Earth Science*, 8(1),
35 93–103. <https://doi.org/10.1007/s11707-013-0390-y>
- 36 Yin, Y., Ma, D., Wu, S., & Pan, T. (2015). Projections of aridity and its regional variability over
37 China in the mid-21st century. *International Journal of Climatology*, 35(14), 4387–4398.
38 <https://doi.org/10.1002/joc.4295>
- 39 Yosef, B. A. A., & Asmamaw, D. K. K. (2015). Rainwater harvesting: An option for dry land
40 agriculture in arid and semi-arid Ethiopia. *International Journal of Water Resources and
41 Environmental Engineering*, 7(2), 17–28. <https://doi.org/10.5897/IJWREE2014.0539>
- 42 You, N., Meng, J., & Zhu, L. (2018). Sensitivity and resilience of ecosystems to climate variability in
43 the semi- arid to hyper-arid areas of Northern China: a case study in the Heihe River Basin.
44 *Ecological Research*, 33(1), 161–174. <https://doi.org/10.1007/s11284-017-1543-3>
- 45 Yu, K., D'Odorico, P., Bhattachan, A., Okin, G. S., & Evan, A. T. (2015). Dust-rainfall feedback in
46 West African Sahel. *Geophysical Research Letters*, 42(18), 7563–7571.
47 <https://doi.org/10.1002/2015GL065533>

- 1 Yu Qiu, G., Lee, I.-B., Shimizu, H., Gao, Y., & Ding, G. (2004). Principles of sand dune fixation with
2 straw checkerboard technology and its effects on the environment. *Journal of Arid*
3 *Environments*, 56(3), 449–464. [https://doi.org/10.1016/S0140-1963\(03\)00066-1](https://doi.org/10.1016/S0140-1963(03)00066-1)
- 4 Yu, Y., Notaro, M., Liu, Z., Kalashnikova, O., Alkolibi, F., Fadda, E., ... Bakhry, F. (2013).
5 Assessing temporal and spatial variations in atmospheric dust over Saudi Arabia through
6 satellite, radiometric, and station data. *J. Geophys. Res. Atmos*, 118, 253–266.
7 <https://doi.org/10.1002/2013JD020677>
- 8 Yurtoglu, M. A. (2015). The Afforestation and Erosion Control Mobilization Action Plan in Turkey.
9 In Dr Jacqui Griffiths (Ed.), *Living Land* (pp. 37–39). UNCCD and Tudor Rose.
- 10 Yusa, A., Berry, P., J.Cheng, J., Ogden, N., Bonsal, B., Stewart, R., ... Waldick, R. (2015). Climate
11 Change, Drought and Human Health in Canada. *International Journal of Environmental*
12 *Research and Public Health*, 12(7), 8359–8412. <https://doi.org/10.3390/ijerph120708359>
- 13 Zabaleta, A., Meaurio, M., Ruiz, E., & Antigüedad, I. (2013). Simulation Climate Change Impact on
14 Runoff and Sediment Yield in a Small Watershed in the Basque Country, Northern Spain.
15 *Journal of Environment Quality*, 43(1), 235. <https://doi.org/10.2134/jeq2012.0209>
- 16 Zaitchik, B. F., Evans, J. P., Geerken, R. A., Smith, R. B., Zaitchik, B. F., Evans, J. P., ... Smith, R.
17 B. (2007). Climate and Vegetation in the Middle East: Interannual Variability and Drought
18 Feedbacks. *Journal of Climate*, 20(15), 3924–3941. <https://doi.org/10.1175/JCLI4223.1>
- 19 Zaitchik, B. F., Santanello, J. A., Kumar, S. V., Peters-Lidard, C. D., Zaitchik, B. F., Santanello, J.
20 A., ... Peters-Lidard, C. D. (2013). Representation of Soil Moisture Feedbacks during Drought
21 in NASA Unified WRF (NU-WRF). *Journal of Hydrometeorology*, 14(1), 360–367.
22 <https://doi.org/10.1175/JHM-D-12-069.1>
- 23 Zare, M., Nazari Samani, A. A., Mohammady, M., Teimurian, T., & Bazrafshan, J. (2016).
24 Simulation of soil erosion under the influence of climate change scenarios. *Environmental Earth*
25 *Sciences*, 75(21), 1–15. <https://doi.org/10.1007/s12665-016-6180-6>
- 26 Zdruli, P. (2011). *Desertification in the Mediterranean Region*. Girona, Barcelona: European Institute
27 of the Mediterranean. Retrieved from [http://www.iemed.org/observatori-en/arees-](http://www.iemed.org/observatori-en/arees-danalisi/arxius-adjunts/anuari/med.2011/Zdrulli_en.pdf)
28 [danalisi/arxius-adjunts/anuari/med.2011/Zdrulli_en.pdf](http://www.iemed.org/observatori-en/arees-danalisi/arxius-adjunts/anuari/med.2011/Zdrulli_en.pdf)
- 29 Zeng, N., & Yoon, J. (2009). Expansion of the world's deserts due to vegetation-albedo feedback
30 under global warming. *Geophysical Research Letters*, 36(17), L17401.
31 <https://doi.org/10.1029/2009GL039699>
- 32 Zhang, J., Zhang, Y., Fan, D., Qin, S., Jia, X., Wu, B., ... Zhu, L. (2015). Effects of Sand Dune
33 Stabilization on the Spatial Pattern of *Artemisia ordosica* Population in Mu Us Desert,
34 Northwest China. *PLOS ONE*, 10(6), e0129728. <https://doi.org/10.1371/journal.pone.0129728>
- 35 Zhang, K., Qu, J., Liao, K., Niu, Q., & Han, Q. (2010). Damage by wind-blown sand and its control
36 along Qinghai-Tibet Railway in China. *Aeolian Research*, 1(3–4), 143–146.
37 <https://doi.org/10.1016/J.AEOLIA.2009.10.001>
- 38 Zhang, W., Zhou, J., Feng, G., Weindorf, D. C., Hu, G., & Sheng, J. (2015). Characteristics of water
39 erosion and conservation practice in arid regions of Central Asia: Xinjiang Province, China as an
40 example. *International Soil and Water Conservation Research*, 3(2), 97–111.
41 <https://doi.org/10.1016/j.iswcr.2015.06.002>
- 42 Zhang, X. C., & Nearing, M. A. (2005). Impact of climate change on soil erosion, runoff, and wheat
43 productivity in central Oklahoma. *CATENA*, 61(2–3), 185–195.
44 <https://doi.org/10.1016/J.CATENA.2005.03.009>
- 45 Zhang, X. Y., Arimoto, R., & An, Z. S. (1997). Dust emission from Chinese desert sources linked to
46 variations in atmospheric circulation. *Journal of Geophysical Research: Atmospheres*,
47 102(D23), 28041–28047. <https://doi.org/10.1029/97JD02300>
- 48 Zhang, X., Zhao, L., Tong, D., Wu, G., Dan, M., Teng, B., ... Teng, B. (2016). A Systematic Review

- 1 of Global Desert Dust and Associated Human Health Effects. *Atmosphere*, 7(12), 158.
2 <https://doi.org/10.3390/atmos7120158>
- 3 Zhang, Y., Song, C., Band, L. E., Sun, G., & Li, J. (2017). Reanalysis of global terrestrial vegetation
4 trends from MODIS products: Browning or greening? *Remote Sensing of Environment*, 191,
5 145–155. <https://doi.org/10.1016/J.RSE.2016.12.018>
- 6 Zhang, Z., & Huisingh, D. (2018). Combating desertification in China: Monitoring, control,
7 management and revegetation. *Journal of Cleaner Production*, 182, 765–775.
8 <https://doi.org/10.1016/J.JCLEPRO.2018.01.233>
- 9 Zhao, C., Liu, B., Piao, S., Wang, X., Lobell, D. B., Huang, Y., ... Asseng, S. (2017). Temperature
10 increase reduces global yields of major crops in four independent estimates. *Proceedings of the*
11 *National Academy of Sciences of the United States of America*, 114(35), 9326–9331.
12 <https://doi.org/10.1073/pnas.1701762114>
- 13 Zhao, R., Chen, Y., Shi, P., Zhang, L., Pan, J., & Zhao, H. (2013). Land use and land cover change
14 and driving mechanism in the arid inland river basin: a case study of Tarim River, Xinjiang,
15 China. *Environmental Earth Sciences*, 68(2), 591–604. [https://doi.org/10.1007/s12665-012-](https://doi.org/10.1007/s12665-012-1763-3)
16 [1763-3](https://doi.org/10.1007/s12665-012-1763-3)
- 17 Zhao, S., Peng, C., Jiang, H., Tian, D., Lei, X., & Zhou, X. (2006). Land use change in Asia and the
18 ecological consequences. *Ecological Research*, 21(6), 890–896. [https://doi.org/10.1007/s11284-](https://doi.org/10.1007/s11284-006-0048-2)
19 [006-0048-2](https://doi.org/10.1007/s11284-006-0048-2)
- 20 Zhao, S., Zhang, H., Feng, S., & Fu, Q. (2015). Simulating direct effects of dust aerosol on arid and
21 semi-arid regions using an aerosol-climate coupled system. *International Journal of*
22 *Climatology*, 35(8), 1858–1866. <https://doi.org/10.1002/joc.4093>
- 23 Zhao, T., & Dai, A. (2015). The Magnitude and Causes of Global Drought Changes in the Twenty-
24 First Century under a Low–Moderate Emissions Scenario. *Journal of Climate*, 28(11), 4490–
25 4512. <https://doi.org/10.1175/JCLI-D-14-00363.1>
- 26 Zheng, J., Yu, Y., Zhang, X., & Hao, Z. (2018). Variation of extreme drought and flood in North
27 China revealed by document-based seasonal precipitation reconstruction for the past 300 years.
28 *Climate of the Past*, 14(8), 1135–1145. <https://doi.org/10.5194/cp-14-1135-2018>
- 29 Zhou, D., Lin, Z., Liu, L., & Zimmermann, D. (2013). Assessing secondary soil salinization risk
30 based on the PSR sustainability framework. *Journal of Environmental Management*, 128, 642–
31 654. <https://doi.org/10.1016/J.JENVMAN.2013.06.025>
- 32 Zhou, W., Gang, C., Zhou, F., Li, J., Dong, X., & Zhao, C. (2015). Quantitative assessment of the
33 individual contribution of climate and human factors to desertification in northwest China using
34 net primary productivity as an indicator. *Ecological Indicators*, 48, 560–569.
35 <https://doi.org/10.1016/J.ECOLIND.2014.08.043>
- 36 Zhu, J., Kang, H., & Xu, M.-L. (2007). Natural regeneration barriers of *Pinus sylvestris* var.
37 *mongolica* plantations in southern Keerqin sandy land, China. *Shengtai Xuebao/ Acta Ecologica*
38 *Sinica*, 27, 4086–4095.
- 39 Zhu, T., & Ringler, C. (2012). Climate Change Impacts on Water Availability and Use in the
40 Limpopo River Basin. *Water*, (4), 63–84.
- 41 Zhu, Z., Piao, S., Myneni, R. B., Huang, M., Zeng, Z., Canadell, J. G., ... Zeng, N. (2016). Greening
42 of the Earth and its drivers. *Nature Climate Change*, 6(8), 791–795.
43 <https://doi.org/10.1038/nclimate3004>
- 44 Ziadat, F. M., & Taimeh, A. Y. (2013). Effect Of Rainfall Intensity, Slope, Land Use And Antecedent
45 Soil Moisture On Soil Erosion In An Arid Environment. *Land Degradation & Development*,
46 24(6), 582–590. <https://doi.org/10.1002/ldr.2239>
- 47 Zidane, I. F., Saqr, K. M., Swadener, G., Ma, X., & Shehadeh, M. F. (2016). On the role of surface
48 roughness in the aerodynamic performance and energy conversion of horizontal wind turbine

- 1 blades: a review. *International Journal of Energy Research*, 40(15), 2054–2077.
2 <https://doi.org/10.1002/er.3580>
- 3 Ziese, M., Schneider, U., Meyer-Christoffer, A., Schamm, K., Vido, J., Finger, P., ... Becker, A.
4 (2014). The GPCC Drought Index – a new, combined and gridded global drought index. *Earth*
5 *System Science Data*, 6(2), 285–295. <https://doi.org/10.5194/essd-6-285-2014>
- 6 Zika, M., & Erb, K. H. (2009). The global loss of net primary production resulting from human-
7 induced soil degradation in drylands. *Ecological Economics*, 69(2), 310–318.
8 <https://doi.org/10.1016/j.ecolecon.2009.06.014>
- 9 Zinabu, S., Kebede, A., Ferede, B., & Dugassa, J. (2018). Climate Change prevalence of animal
10 diseases, 4(15).
- 11 Zou, X. K., & Zhai, P. M. (2004). Relationship between vegetation coverage and spring dust storms
12 over northern China. *Journal of Geophysical Research: Atmospheres*, 109(D3), n/a-n/a.
13 <https://doi.org/10.1029/2003JD003913>
- 14 Zucca, C., Peruta, R. Della, Salvia, R., Sommer, S., & Cherlet, M. (2012). Towards a World
15 Desertification Atlas. Relating and selecting indicators and data sets to represent complex issues.
16 *Ecological Indicators*, 15(1), 157–170. <https://doi.org/10.1016/J.ECOLIND.2011.09.012>
- 17 Zulu, L. C. (2010). The forbidden fuel: Charcoal, urban woodfuel demand and supply dynamics,
18 community forest management and woodfuel policy in Malawi. *Energy Policy*, 38(7), 3717–
19 3730. <https://doi.org/10.1016/J.ENPOL.2010.02.050>
- 20 Tokbergenova, A., Nyussupova, G., Arslan, M., & Kiyassova, S. K. L. (2018). Causes and Impacts of
21 Land Degradation and Desertification: Case Study from Kazakhstan. In *Vegetation of Central*
22 *Asia and Environs* (pp. 291–302). Cham: Springer International Publishing.
23 https://doi.org/10.1007/978-3-319-99728-5_11
- 24

1 **Chapter 4: Land Degradation**

2

3 **Coordinating Lead Authors:** Lennart Olsson (Sweden), Humberto Barbosa (Brazil)

4 **Lead Authors:** Suruchi Bhadwal (India), Annette Cowie (Australia), Kenel Delusca (Haiti), Dulce
5 Flores-Renteria (Mexico), Kathleen Hermans (Germany), Esteban Jobbagy (Argentina), Werner Kurz
6 (Canada), Diqiang Li (China), Denis Jean Sonwa (Cameroon), Lindsay Stringer (United Kingdom)

7 **Contributing Authors:** Timothy Crews (The United States of America), Martin Dallimer (United
8 Kingdom), Joris Eekhout (The Netherlands), Karlheinz Erb (Italy), Eamon Haughey (Ireland),
9 Richard Houghton (The United States of America), Muhammad Mohsin Iqbal (Pakistan), Francis X.
10 Johnson (The United States of America), Woo-Kyun Lee (The Republic of Korea), John Morton
11 (United Kingdom), Felipe Garcia Oliva (Mexico), Jan Petzold (Germany), Mohammad Rahimi (Iran),
12 Florence Renou-Wilson (Ireland), Anna Tengberg (Sweden), Louis Verchot (Colombia/The United
13 States of America), Katharine Vincent (South Africa)

14 **Review Editors:** José Manuel Moreno Rodriguez (Spain), Carolina Vera (Argentina)

15 **Chapter Scientist:** Aliyu Salisu Barau (Nigeria)

16 **Date of Draft:** 07/08/2019

17

1

2 **Table of Contents**

| | | |
|----|--|------|
| 3 | Chapter 4: Land Degradation..... | 4-1 |
| 4 | Executive Summary | 4-4 |
| 5 | 4.1 Introduction..... | 4-8 |
| 6 | 4.1.1 Scope of the chapter..... | 4-8 |
| 7 | 4.1.2 Perspectives of land degradation..... | 4-8 |
| 8 | 4.1.3 Definition of land degradation | 4-9 |
| 9 | 4.1.4 Land degradation in previous IPCC reports..... | 4-10 |
| 10 | 4.1.5 Sustainable land management and sustainable forest management | 4-11 |
| 11 | 4.1.6 The human dimension of land degradation and forest degradation | 4-14 |
| 12 | 4.2 Land degradation in the context of climate change | 4-15 |
| 13 | 4.2.1 Processes of land degradation | 4-16 |
| 14 | 4.2.2 Drivers of land degradation..... | 4-25 |
| 15 | 4.2.3 Attribution in the case of land degradation | 4-26 |
| 16 | 4.2.4 Approaches to assessing land degradation | 4-30 |
| 17 | 4.3 Status and current trends of land degradation | 4-33 |
| 18 | 4.3.1 Land degradation..... | 4-33 |
| 19 | 4.3.2 Forest degradation..... | 4-36 |
| 20 | 4.4 Projections of land degradation in a changing climate | 4-40 |
| 21 | 4.4.1 Direct impacts on land degradation..... | 4-40 |
| 22 | 4.4.2 Indirect impacts on land degradation | 4-45 |
| 23 | 4.5 Impacts of bioenergy and technologies for CO ₂ removal (CDR) on land degradation..... | 4-46 |
| 24 | 4.5.1 Potential scale of bioenergy and land-based CDR..... | 4-46 |
| 25 | 4.5.2 Risks of land degradation from expansion of bioenergy and land-based CDR | 4-46 |
| 26 | 4.5.3 Potential contributions of land-based CDR to reducing and reversing land degradation | |
| 27 | 4-47 | |
| 28 | 4.5.4 Traditional biomass provision and land degradation | 4-48 |
| 29 | 4.6 Impacts of land degradation on climate | 4-49 |
| 30 | 4.6.1 Impacts on greenhouse gases | 4-50 |
| 31 | 4.6.2 Physical impacts..... | 4-51 |
| 32 | 4.7 Impacts of climate-related land degradation on poverty and livelihoods | 4-53 |
| 33 | 4.7.1 Relationships between land degradation, climate change and poverty | 4-53 |
| 34 | 4.7.2 Impacts of climate related land degradation on food security | 4-56 |
| 35 | 4.7.3 Impacts of climate-related land degradation on migration and conflict..... | 4-57 |
| 36 | 4.8 Addressing land degradation in the context of climate change..... | 4-58 |

| | | | |
|----|-------|---|------|
| 1 | 4.8.1 | Actions on the ground to address land degradation | 4-59 |
| 2 | 4.8.2 | Local and indigenous knowledge for addressing land degradation | 4-63 |
| 3 | 4.8.3 | Reducing deforestation and forest degradation and increasing afforestation..... | 4-64 |
| 4 | 4.8.4 | Sustainable forest management and CO ₂ removal technologies | 4-65 |
| 5 | 4.8.5 | Policy responses to land degradation | 4-67 |
| 6 | 4.8.6 | Resilience and thresholds..... | 4-69 |
| 7 | 4.8.7 | Barriers to implementation of sustainable land management | 4-71 |
| 8 | 4.9 | Case-studies | 4-73 |
| 9 | 4.9.1 | Urban green infrastructure | 4-74 |
| 10 | 4.9.2 | Perennial Grains and Soil Organic Carbon | 4-76 |
| 11 | 4.9.3 | Reversing land degradation through reforestation | 4-79 |
| 12 | 4.9.4 | Degradation and management of peat soils | 4-82 |
| 13 | 4.9.5 | Biochar..... | 4-84 |
| 14 | 4.9.6 | Management of land degradation induced by tropical cyclones | 4-87 |
| 15 | 4.9.7 | Saltwater intrusion | 4-89 |
| 16 | 4.9.8 | Avoiding coastal maladaptation..... | 4-91 |
| 17 | 4.10 | Knowledge gaps and key uncertainties | 4-92 |
| 18 | | Frequently Asked Questions | 4-93 |
| 19 | | References..... | 4-94 |
| 20 | | | |
| 21 | | | |

1 **Executive Summary**

2 **Land degradation affects people and ecosystems throughout the planet and is both affected by**
3 **climate change and contributes to it.** In this report, land degradation is defined as a *negative trend*
4 *in land condition, caused by direct or indirect human-induced processes including anthropogenic*
5 *climate change, expressed as long-term reduction or loss of at least one of the following: biological*
6 *productivity, ecological integrity, or value to humans.* Forest degradation is land degradation which
7 occurs in forest land. Deforestation is the conversion of forest to non-forest land and can result in land
8 degradation. {4.1.3}

9 **Land degradation adversely affects people's livelihoods (*very high confidence*) and occurs over a**
10 **quarter of the Earth's ice-free land area (*medium confidence*).** **The majority of the 1.3 to 3.2**
11 **billion affected people (*low confidence*) are living in poverty in developing countries (*medium***
12 ***confidence*).** Land use changes and unsustainable land management are direct human causes of land
13 degradation (*very high confidence*), with agriculture being a dominant sector driving degradation
14 (*very high confidence*). Soil loss from conventionally tilled land exceeds the rate of soil formation by
15 >2 orders of magnitude (*medium confidence*). Land degradation affects humans in multiple ways,
16 interacting with social, political, cultural and economic aspects, including markets, technology,
17 inequality and demographic change (*very high confidence*). Land degradation impacts extend beyond
18 the land surface itself, affecting marine and freshwater systems, as well as people and ecosystems far
19 away from the local sites of degradation (*very high confidence*). {4.1.6, 4.2.1, 4.2.3, 4.3, 4.6.1, 4.7,
20 Table 4.1}

21 **Climate change exacerbates the rate and magnitude of several ongoing land degradation**
22 **processes and introduces new degradation patterns (*high confidence*).** Human-induced global
23 warming has already caused observed changes in two drivers of land degradation: increased
24 frequency, intensity and/or amount of heavy precipitation (*medium confidence*), and increased heat
25 stress (*high confidence*). Global warming beyond that of present-day will further exacerbate ongoing
26 land degradation processes through increasing floods (*medium confidence*), drought frequency and
27 severity (*medium confidence*), intensified cyclones (*medium confidence*), and sea-level rise (*very high*
28 *confidence*), with outcomes being modulated by land management (*very high confidence*). Permafrost
29 thawing due to warming (*high confidence*), and coastal erosion due to sea level rise and impacts of
30 changing storm paths (*low confidence*), are examples of land degradation affecting places in which it
31 has not typically been a problem. Erosion of coastal areas because of sea level rise will increase
32 worldwide (*high confidence*). In cyclone prone areas the combination of sea level rise and more
33 intense cyclones will cause land degradation with serious consequences for people and livelihoods
34 (*very high confidence*). {4.2.1, 4.2.2, 4.2.3, 4.4.1, 4.4.2, 4.9.6, Table 4.1}

35 **Land degradation and climate change, both individually and in combination, have profound**
36 **implications for natural resource-based livelihood systems and societal groups (*high***
37 ***confidence*).** The number of people whose livelihood depends on degraded lands has been estimated
38 to ~1.5 billion worldwide (*very low confidence*). People in degraded areas who directly depend on
39 natural resources for subsistence, food security and income, including women and youth with limited
40 adaptation options, are especially vulnerable to land degradation and climate change (*high*
41 *confidence*). Land degradation reduces land productivity and increases the workload of managing the
42 land, affecting women disproportionately in some regions. Land degradation and climate change act as
43 threat multipliers for already precarious livelihoods (*very high confidence*), leaving them highly
44 sensitive to extreme climatic events, with consequences such as poverty and food insecurity (*high*
45 *confidence*), and in some cases migration, conflict and loss of cultural heritage (*low confidence*).
46 Changes in vegetation cover and distribution due to climate change increase risks of land degradation
47 in some areas (*medium confidence*). Climate change will have detrimental effects on livelihoods,

1 habitats, and infrastructure through increased rates of land degradation (*high confidence*) and from
2 new degradation patterns (*low evidence, high agreement*). {4.1.6, 4.2.1, 4.7}

3 **Land degradation is a driver of climate change through emission of greenhouse gases and**
4 **reduced rates of carbon uptake (*very high confidence*).** Since 1990, globally the forest area has
5 decreased by 3% (*low confidence*) with net decreases in the tropics and net increases outside the
6 tropics (*high confidence*). Lower carbon density in re-growing forests compared to carbon stocks
7 before deforestation results in net emissions from land use change (*very high confidence*). Forest
8 management that reduces carbon stocks of forest land also leads to emissions, but global estimates of
9 these emissions are uncertain. Cropland soils have lost 20-60% of their organic carbon content prior
10 to cultivation, and soils under conventional agriculture continue to be a source of greenhouse gases
11 (*medium confidence*). Of the land degradation processes, deforestation, increasing wildfires,
12 degradation of peat soils, and permafrost thawing contribute most to climate change through the
13 release of greenhouse gases and the reduction in land carbon sinks following deforestation (*high*
14 *confidence*). Agricultural practices also emit non-CO₂ greenhouse gases from soils and these
15 emissions are exacerbated by climate change (*medium confidence*). Conversion of primary to
16 managed forests, illegal logging and unsustainable forest management result in greenhouse gas
17 emissions (*very high confidence*) and can have additional physical effects on the regional climate
18 including those arising from albedo shifts (*medium confidence*). These interactions call for more
19 integrative climate impact assessments. {4.2.2, 4.3, 4.5.4, 4.6}

20 **Large-scale implementation of dedicated biomass production for bioenergy increases**
21 **competition for land with potentially serious consequences for food security and land**
22 **degradation (*high confidence*).** Increasing the extent and intensity of biomass production through
23 e.g. fertiliser additions, irrigation or monoculture energy plantations can result in local land
24 degradation. Poorly implemented intensification of land management contributes to land degradation
25 (e.g., salinisation from irrigation) and disrupted livelihoods (*high confidence*). In areas where
26 afforestation and reforestation occur on previously degraded lands, opportunities exist to restore and
27 rehabilitate lands with potentially significant co-benefits (*high confidence*) that depend on whether
28 restoration involves natural or plantation forests. The total area of degraded lands has been estimated
29 at 1-6 Mkm² (*very low confidence*). The extent of degraded and marginal lands suitable for dedicated
30 biomass production is highly uncertain and cannot be established without due consideration of current
31 land use and land tenure. Increasing the area of dedicated energy crops can lead to land degradation
32 elsewhere through indirect land use change (*medium confidence*). Impacts of energy crops can be
33 reduced through strategic integration with agricultural and forestry systems (*high confidence*) but the
34 total quantity of biomass that can be produced through synergistic production systems is unknown.
35 {4.1.6, 4.4.2, 4.5, 4.7.1, 4.8.1, 4.8.3, 4.8.4, 4.9.3}

36 **Reducing unsustainable use of traditional biomass reduces land degradation and emissions of**
37 **CO₂, while providing social and economic co-benefits (*very high confidence*).** Traditional biomass
38 in the form of fuelwood, charcoal and agricultural residues remains a primary source of energy for
39 more than one-third of the global population leading to unsustainable use of biomass resources and
40 forest degradation and contributing around 2% of global greenhouse gas (GHG) emissions (*low*
41 *confidence*). Enhanced forest protection, improved forest and agricultural management, fuel-switching
42 and adoption of efficient cooking and heating appliances can promote more sustainable biomass use
43 and reduce land degradation, with co-benefits of reduced GHG emissions, improved human health,
44 and reduced workload especially for women and youth (*very high confidence*). {4.1.6, 4.5.4}

45 **Land degradation can be avoided, reduced or reversed by implementing sustainable land**
46 **management, restoration and rehabilitation practices that simultaneously provide many co-**
47 **benefits, including adaptation to and mitigation of climate change (*high confidence*).** Sustainable

1 land management is a comprehensive array of technologies and enabling conditions, which have
2 proven to address land degradation at multiple landscape scales, from local farms (*very high*
3 *confidence*) to entire watersheds (*medium confidence*). Sustainable forest management can prevent
4 deforestation, maintain and enhance carbon sinks and can contribute towards greenhouse gas
5 emissions reduction goals. Sustainable forest management generates socio-economic benefits,
6 provides fiber, timber and biomass to meet society's growing needs. While sustainable forest
7 management sustains high carbon sinks, the conversion from primary forests to sustainably managed
8 forests can result in carbon emission during the transition and can result in loss of biodiversity (*high*
9 *confidence*). Conversely, in areas of degraded forests, sustainable forest management can increase
10 carbon stocks and biodiversity (*medium confidence*). Carbon storage in long-lived wood products and
11 reductions of emissions from use of wood products to substitute for emissions-intensive materials also
12 contribute to mitigation objectives. {4.8, 4.9, Table 4.2}

13 **Lack of action to address land degradation will increase emissions and reduce carbon sinks and**
14 **is inconsistent with the emission reductions required to limit global warming to 1.5°C or 2°C.**
15 **(*high confidence*).** Better management of soils can offset 5–20% of current global anthropogenic
16 GHG emissions (*medium confidence*). Measures to avoid, reduce and reverse land degradation are
17 available but economic, political, institutional, legal and socio-cultural barriers, including lack of
18 access to resources and knowledge, restrict their uptake (*very high confidence*). Proven measures that
19 facilitate implementation of practices that avoid, reduce, or reverse land degradation include tenure
20 reform, tax incentives, payments for ecosystem services, participatory integrated land use planning,
21 farmer networks and rural advisory services. Delayed action increases the costs of addressing land
22 degradation, and can lead to irreversible biophysical and human outcomes (*high confidence*). Early
23 actions can generate both site specific and immediate benefits to communities affected by land
24 degradation, and contribute to long-term global benefits through climate change mitigation (*high*
25 *confidence*). {4.1.5, 4.1.6, 4.7.1, 4.8, Table 4.2}

26 **Even with adequate implementation of measures to avoid, reduce and reverse land degradation**
27 **there will be residual degradation in some situations (*high confidence*).** Limits to adaptation are
28 dynamic, site specific and are determined through the interaction of biophysical changes with social
29 and institutional conditions. Exceeding the limits of adaptation will trigger escalating losses or result
30 in undesirable changes, such as forced migration, conflicts, or poverty. Examples of potential limits to
31 adaptation due to climate change induced land degradation are coastal erosion where land disappears,
32 collapsing infrastructure and livelihoods due to thawing of permafrost, and extreme forms of soil
33 erosion. {4.7, 4.8.5, 4.8.6, 4.9.6, 4.9.7, 4.9.8}

34 **Land degradation is a serious and widespread problem, yet key uncertainties remain**
35 **concerning its extent, severity, and linkages to climate change (*very high confidence*).** Despite the
36 difficulties of objectively measuring the extent and severity of land degradation given its complex and
37 value-based characteristics, land degradation represents, like climate change, one of the biggest and
38 most urgent challenges for humanity (*very high confidence*). The current global extent, severity and
39 rates of land degradation are not well quantified. There is no single method by which land degradation
40 can be measured objectively and consistently over large areas because it is such a complex and value
41 laden concept (*very high confidence*). However, many scientific and locally-based approaches,
42 including the use of indigenous and local knowledge, exist that can assess different aspects of land
43 degradation or provide proxies. Remote sensing, corroborated by other data, can generate
44 geographically explicit and globally consistent data that can be used as proxies over relevant time
45 scales (several decades). Few studies have specifically addressed the impacts of proposed land-based
46 negative emission technologies on land degradation. Much research has tried to understand how
47 livelihoods and ecosystems are affected by a particular stressor, for example drought, heat stress, or
48 water logging. Important knowledge gaps remain in understanding how plants, habitats and

- 1 ecosystems are affected by the cumulative and interacting impacts of several stressors, including
- 2 potential new stressors resulting from large-scale implementation of negative emission technologies.
- 3 {4.10}
- 4

1 **4.1 Introduction**

2 **4.1.1 Scope of the chapter**

3 This chapter examines the scientific understanding of how climate change impacts land degradation,
4 and vice versa, with a focus on non-drylands. Land degradation of drylands is covered in Chapter 3.
5 After providing definitions and the context (Section 4.1) we proceed with a theoretical explanation of
6 the different processes of land degradation and how they are related to climate and to climate change,
7 where possible (Section 4.2). Two sections are devoted to a systematic assessment of the scientific
8 literature on status and trend of land degradation (Section 4.3) and projections of land degradation
9 (Section 4.4). Then follows a section where we assess the impacts of climate change mitigation
10 options, bioenergy and land-based technologies for carbon dioxide removal (CDR), on land
11 degradation (Section 4.5). The ways in which land degradation can impact climate and climate change
12 are assessed in Section 4.6. The impacts of climate related land degradation on human and natural
13 systems are assessed in Section 4.7. The remainder of the chapter assesses land degradation mitigation
14 options based on the concept of sustainable land management: avoid, reduce and reverse land
15 degradation (Section 4.8), followed by a presentation of eight illustrative case studies of land
16 degradation and remedies (Section 4.9). The chapter ends with a discussion of the most critical
17 knowledge gaps and areas for further research (Section 4.10).

18 **4.1.2 Perspectives of land degradation**

19 Land degradation has accompanied humanity at least since the widespread adoption of agriculture
20 during Neolithic time, some 10,000 to 7,500 years ago (Dotterweich 2013; Butzer 2005; Dotterweich
21 2008) and the associated population increase (Bocquet-Appel 2011). There are indications that the
22 levels of greenhouse gases (particularly carbon dioxide and methane) of the atmosphere started to
23 increase already more than 3,000 years ago as a result of expanding agriculture, clearing of forests,
24 and domestication of wild animals (Fuller et al. 2011; Kaplan et al. 2011; Vavrus et al. 2018; Ellis et
25 al. 2013). While the development of agriculture (cropping and animal husbandry) underpinned the
26 development of civilisations, political institutions, and prosperity, farming practices led to conversion
27 of forests and grasslands to farmland, and the heavy reliance on domesticated annual grasses for our
28 food production meant that soils started to deteriorate through seasonal mechanical disturbances
29 (Turner et al. 1990; Steffen et al. 2005; Ojima et al. 1994; Ellis et al. 2013). More recently,
30 urbanisation has significantly altered ecosystems, see further Cross-chapter Box 4 on Climate Change
31 and Urbanisation, Chapter 2. Since about 1850, about 35% of the human caused emissions of CO₂ to
32 the atmosphere comes from land as a combined effect of land degradation and land-use change (Foley
33 et al. 2005) and about 38% of Earth's land area has been converted to agriculture (Foley et al. 2011),
34 see Chapter 2 for more details.

35 Not all human impacts on land result in degradation according to the definition of land degradation
36 used in this report (see 4.2.1). There are many examples of long-term sustainably managed land
37 around the world (such as terraced agricultural systems and sustainably managed forests) although
38 degradation and its management are the focus of this chapter. We also acknowledge that human use of
39 land and ecosystems provides essential goods and services for society (Foley et al. 2005; MEA
40 (Millennium Ecosystem Assessment) 2005).

41 Land degradation was long subject to a polarised scientific debate between disciplines and
42 perspectives in which social scientists often proposed that natural scientists exaggerated land
43 degradation as a global problem (Blaikie and Brookfield 1987; Forsyth 1996; Lukas 2014; Zimmerer
44 1993). The elusiveness of the concept in combination with the difficulties of measuring and
45 monitoring land degradation at global and regional scales by extrapolation and aggregation of
46 empirical studies at local scales, such as the Global Assessment of Soil Degradation database

1 (GLASOD) (Sonneveld and Dent 2009) contributed to conflicting views. The conflicting views were
2 not confined to science only, but also caused tension between the scientific understanding of land
3 degradation and policy (Andersson et al. 2011; Behnke and Mortimore 2016; Grainger 2009; Toulmin
4 and Brock 2016). Another weakness of many land degradation studies is the exclusion of the views
5 and experiences of the land users, whether farmers or forest dependent communities (Blaikie and
6 Brookfield 1987; Fairhead and Scoones 2005; Warren 2002; Andersson et al. 2011). More recently,
7 the polarised views described above have been reconciled under the umbrella of Land Change
8 Science, which has emerged as an interdisciplinary field aimed at examining the dynamics of land
9 cover and land-use as a coupled human–environment system (Turner et al. 2007). A comprehensive
10 discussion about concepts and different perspectives of land degradation was presented in Chapter 2
11 of the recent report from the Intergovernmental Platform on Biodiversity and Ecosystem Services
12 (IPBES) on land degradation (Montanarella et al. 2018).

13 In summary, agriculture and clearing of land for food and wood products have been the main drivers
14 of land degradation for millennia (*high confidence*). This does not mean, however, that agriculture and
15 forestry always cause land degradation (*high confidence*); sustainable management is possible but not
16 always practiced (*high confidence*). Reasons for this are primarily economic, political and social.

17 **4.1.3 Definition of land degradation**

18 To clarify the scope of this chapter it is important to start by defining land itself. The Special Report
19 on Climate Change and Land (SRCCL) defines land as “the terrestrial portion of the biosphere that
20 comprises the natural resources (soil, near surface air, vegetation and other biota, and water), the
21 ecological processes, topography, and human settlements and infrastructure that operate within that
22 system” (Henry et al. 2018), adapted from (FAO 2007; UNCCD 1994).

23 Land degradation is defined in many different ways within the literature, with differing emphases on
24 biodiversity, ecosystem functions and ecosystem services (e.g., Montanarella et al. 2018). In this
25 report, land degradation is defined as a *negative trend in land condition, caused by direct or indirect*
26 *human-induced processes including anthropogenic climate change, expressed as long-term reduction*
27 *or loss of at least one of the following: biological productivity, ecological integrity or value to*
28 *humans*. This definition applies to forest and non-forest land: forest degradation is land degradation
29 that occurs in forest land. Soil degradation refers to a subset of land degradation processes that
30 directly affect soil.

31 The SRCCL definition is derived from the IPCC AR5 definition of desertification, which is in turn
32 taken from the UNCCD: “Land degradation in arid, semi-arid, and dry sub-humid areas resulting from
33 various factors, including climatic variations and human activities. Land degradation in arid, semi-
34 arid, and dry sub-humid areas is a reduction or loss of the biological or economic productivity and
35 integrity of rainfed cropland, irrigated cropland, or range, pasture, forest, and woodlands resulting
36 from land uses or from a process or combination of processes, including processes arising from
37 human activities and habitation patterns, such as (1) soil erosion caused by wind and/or water; (2)
38 deterioration of the physical, chemical, biological, or economic properties of soil; and (3) long-term
39 loss of natural vegetation” (UNCCD 1994, Article 1).

40 The SRCCL definition is intended to complement the more detailed UNCCD definition, expanding
41 the scope to all regions, not just drylands, providing an operational definition that emphasises the
42 relationship between land degradation and climate for use in this report. Through its attention to the
43 three aspects biological productivity, ecological integrity and value to humans, the SRCCL definition
44 is consistent with the Land Degradation Neutrality (LDN) concept, which aims to maintain or enhance
45 the land-based natural capital, and the ecosystem services that flow from it (Cowie et al. 2018).

1 In the SRCCL definition of land degradation, changes in land condition resulting solely from natural
2 processes (such as volcanic eruptions and tsunamis) are not considered land degradation, as these are
3 not direct or indirect human-induced processes. Climate variability exacerbated by human-induced
4 climate change can contribute to land degradation. Value to humans can be expressed in terms of
5 ecosystem services or Nature's Contribution to People.

6 The definition recognises the reality presented in the literature that land-use and land management
7 decisions often result in trade-offs between time, space, ecosystem services, and stakeholder groups
8 (e.g. Dallimer and Stringer 2018). The interpretation of a negative trend in land condition is somewhat
9 subjective, especially where there is a trade-off between ecological integrity and value to humans. The
10 definition also does not consider the magnitude of the negative trend or the possibility that a negative
11 trend in one criterion may be an acceptable trade-off for a positive trend in another criterion. For
12 example, reducing timber yields to safeguard biodiversity by leaving on site more wood that can
13 provide habitat, or vice versa, is a trade-off that needs to be evaluated based on context (i.e. the
14 broader landscape) and society's priorities. Reduction of biological productivity *or* ecological
15 integrity *or* value to humans *can* constitute degradation, but any one of these changes need not
16 necessarily be considered degradation. Thus, a land-use change that reduces ecological integrity and
17 enhances **sustainable** food production at a specific location is not necessarily degradation. Different
18 stakeholder groups with different world views value ecosystem services differently. As Warren (2002)
19 explained: land degradation is contextual. Further, a decline in biomass carbon stock does not always
20 signify degradation, such as when caused by periodic forest harvest. Even a decline in productivity
21 may not equate to land degradation, such as when a high intensity agricultural system is converted to
22 a lower input more sustainable production system.

23 In the SRCCL definition, degradation is indicated by a negative trend in land condition during the
24 period of interest, thus the baseline is the land condition at the start of this period. The concept of
25 baseline is theoretically important but often practically difficult to implement for conceptual and
26 methodological reasons (Herrick et al. 2019; Prince et al. 2018; see also Sections 4.3.1 and 4.4.1).
27 Especially in biomes characterised by seasonal and interannual variability, the baseline values of the
28 indicators to be assessed should be determined by averaging data over a number of years prior to the
29 commencement of the assessment period (Orr et al. 2017; see also 4.2.4).

30 Forest degradation is land degradation in forest remaining forest. In contrast, deforestation refers to
31 the conversion of forest to non-forest that involves a loss of tree cover and a change in land-use.
32 Internationally accepted definitions of forest (FAO 2015; UNFCCC 2013) include lands where tree
33 cover has been lost temporarily, due to disturbance or harvest, with an expectation of forest regrowth.
34 Such temporary loss of forest cover therefore is not deforestation.

35 **4.1.4 Land degradation in previous IPCC reports**

36 Several previous IPCC assessment reports include brief discussions of land degradation. In AR5
37 WGIII land degradation is one factor contributing to uncertainties of the mitigation potential of land-
38 based ecosystems, particularly in terms of fluxes of soil carbon (Smith et al., 2014, p. 817). In AR5
39 WGI, soil carbon was discussed comprehensively but not in the context of land degradation, except
40 forest degradation (Ciais et al. 2013) and permafrost degradation (Vaughan et al. 2013). Climate
41 change impacts were discussed comprehensively in AR5 WGII, but land degradation was not
42 prominent. Land use and land cover changes were treated comprehensively in terms of effects on the
43 terrestrial carbon stocks and flows (Settele et al. 2015) but links to land degradation were to a large
44 extent missing. Land degradation was discussed in relation to human security as one factor which in
45 combination with extreme weather events has been proposed to contribute to human migration (Adger
46 et al. 2014), an issue discussed more comprehensively in this chapter (see section 4.7.3). Drivers and
47 processes of degradation by which land-based carbon is released to the atmosphere and/or the long-

1 term reduction in the capacity of the land to remove atmospheric carbon and to store this in biomass
2 and soil carbon, have been discussed in the methodological reports of IPCC (IPCC 2006, 2014a) but
3 less so in the assessment reports.

4 The Special Report on Land Use, Land-Use Change and Forestry (SR-LULUCF) (Watson et al. 2000)
5 focused on the role of the biosphere in the global cycles of greenhouse gases (GHG). Land
6 degradation was not addressed in a comprehensive way. Soil erosion was discussed as a process by
7 which soil carbon is lost and the productivity of the land is reduced. Deposition of eroded soil carbon
8 in marine sediments was also mentioned as a possible mechanism for permanent sequestration of
9 terrestrial carbon (Watson et al. 2000) (p. 194). The possible impacts of climate change on land
10 productivity and degradation were not discussed comprehensively. Much of the report was about how
11 to account for sources and sinks of terrestrial carbon under the Kyoto Protocol.

12 The IPCC Special Report on Managing the Risks of Extreme Events and Disasters to Advance
13 Climate Change Adaptation (SREX) (IPCC 2012) did not provide a definition of land degradation.
14 Nevertheless, it addressed different aspects related to some types of land degradation in the context of
15 weather and climate extreme events. From this perspective, it provided key information on both
16 observed and projected changes in weather and climate (extremes) events that are relevant to extreme
17 impacts on socio-economic systems and on the physical components of the environment, notably on
18 permafrost in mountainous areas and coastal zones for different geographic regions, but little explicit
19 links to land degradation. The report also presented the concept of sustainable land management as an
20 effective risk reduction tool.

21 Land degradation has been treated in several previous IPCC reports but mainly as an aggregated
22 concept associated with emissions of GHG or as an issue that can be addressed through adaptation
23 and mitigation.

24 **4.1.5 Sustainable land management and sustainable forest management**

25 Sustainable land management (SLM) is defined as “the stewardship and use of land resources,
26 including soils, water, animals and plants, to meet changing human needs, while simultaneously
27 ensuring the long-term productive potential of these resources and the maintenance of their
28 environmental functions” (Adapted from World Overview of Conservation Approaches and
29 Technologies, WOCAT). Achieving the objective of ensuring that productive potential is maintained
30 in the long term will require implementation of adaptive management and “triple loop learning”, that
31 seeks to monitor outcomes, learn from experience and emerging new knowledge, modifying
32 management accordingly (Rist et al. 2013).

33 Sustainable Forest Management (SFM) is defined as “the stewardship and use of forests and forest
34 lands in a way, and at a rate, that maintains their biodiversity, productivity, regeneration capacity,
35 vitality and their potential to fulfill, now and in the future, relevant ecological, economic and social
36 functions, at local, national, and global levels, and that does not cause damage to other ecosystems”
37 (Forest Europe 2016; Mackey et al. 2015). This SFM definition was developed by the Ministerial
38 Conference on the Protection of Forests in Europe and has since been adopted by the Food and
39 Agriculture Organization. Forest management that fails to meet these sustainability criteria can
40 contribute to land degradation. Land degradation can be reversed through restoration and
41 rehabilitation, which are defined in the Glossary, where other terms that are used but not explicitly
42 defined in this section can also be found. While the definitions of SLM and SFM are very similar and
43 could be merged, both are included to maintain the subtle differences in the existing definitions.

44 Climate change impacts interact with land management to determine sustainable or degraded outcome
45 (Figure 4.1). Climate change can exacerbate many degradation processes (Table 4.1) and introduce
46 novel ones (e.g., permafrost thawing or biome shifts). To avoid, reduce or reverse degradation, land

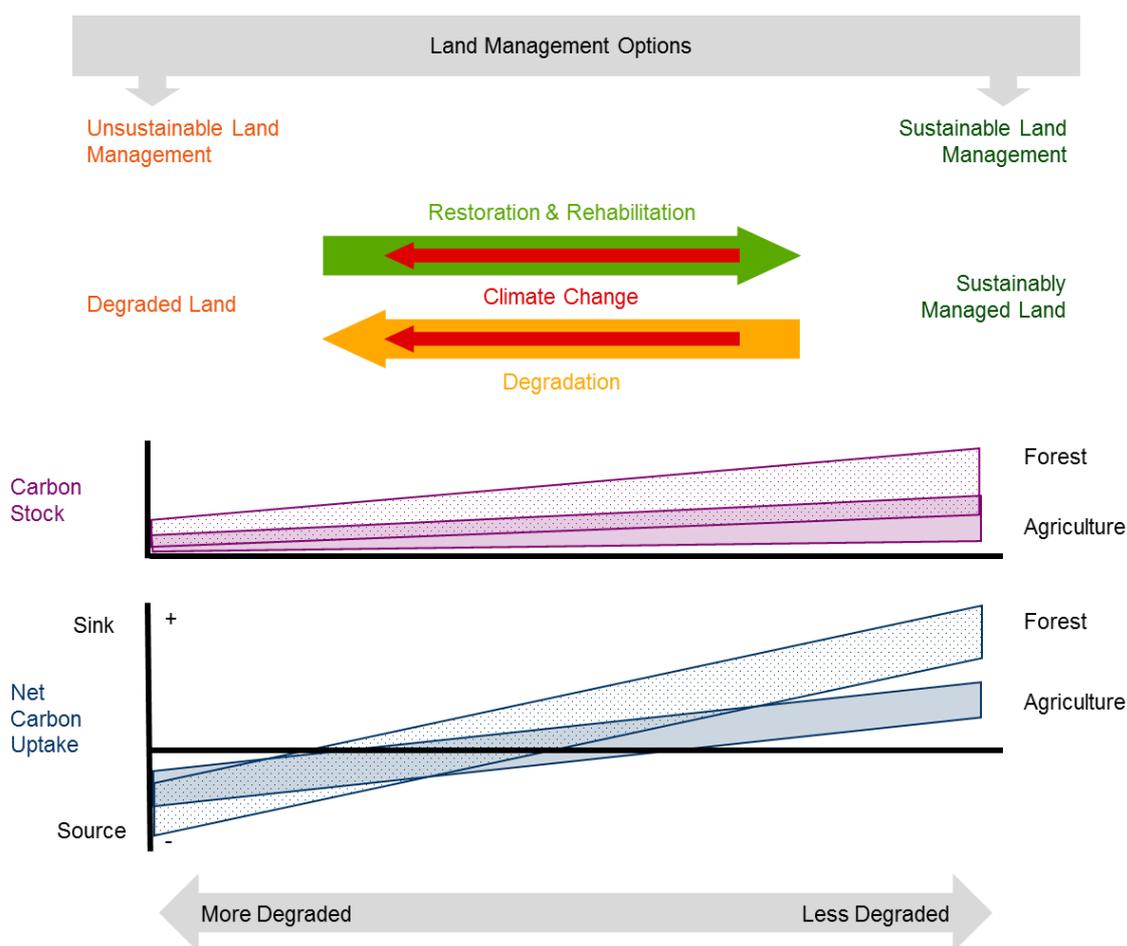
1 management activities can be selected to mitigate the impact of, and adapt to, climate change. In some
2 cases, climate change impacts may result in increased productivity and carbon stocks, at least in the
3 short term. For example, longer growing seasons due to climate warming can lead to higher forest
4 productivity (Henttonen et al. 2017; Kauppi et al. 2014; Dragoni et al. 2011), but warming alone
5 many not increase productivity where other factors such a water supply are limiting (Hember et al.
6 2017).

7 The types and intensity of human land-use and climate change impacts on lands affect their carbon
8 stocks and their ability to operate as carbon sinks. In managed agricultural lands, degradation can
9 result in reductions of soil organic carbon stocks, which also adversely affects land productivity and
10 carbon sinks (See Figure 4.1).

11 The transition from natural to managed forest landscapes usually results in an initial reduction of
12 landscape-level carbon stocks. The magnitude of this reduction is a function of the differential in
13 frequency of stand replacing natural disturbances (e.g. wildfires) and harvest disturbances, as well as
14 the age-dependence of these disturbances (Harmon et al. 1990; Kurz et al. 1998a; Trofymow et al.
15 2008).

16 Sustainable forest management applied at the landscape scale to existing unmanaged forests can first
17 reduce average forest carbon stocks and subsequently increase the rate at which carbon dioxide is
18 removed from the atmosphere, because net ecosystem production of forest stands is highest in
19 intermediate stand ages (Kurz et al. 2013; Volkova et al. 2018; Tang et al. 2014). The net impact on
20 the atmosphere depends on the magnitude of the reduction in carbon stocks, the fate of the harvested
21 biomass (i.e. use in short or long-lived products and for bioenergy, and therefore displacement of
22 emissions associated with GHG-intensive building materials and fossil fuels), and the rate of
23 regrowth. Thus, the impacts of sustainable forest management on one indicator (e.g., past reduction
24 in C stocks in the forested landscape) can be negative, while those on another indicator (e.g., current
25 forest productivity and rate of CO₂ removal from the atmosphere, avoided fossil fuel emissions) can
26 be positive. Sustainably managed forest landscapes can have a lower biomass carbon density than
27 unmanaged forest, but the younger forests can have a higher growth rate, and therefore contribute
28 stronger carbon sinks, than older forests (Trofymow et al. 2008; Volkova et al. 2018; Poorter et al.
29 2016).

30



1
2 **Figure 4.1** Conceptual figure illustrating that climate change impacts interact with land management to
3 determine sustainable or degraded outcome. Climate change can exacerbate many degradation processes
4 (Table 4.1) and introduce novel ones (e.g., permafrost thawing or biome shifts), hence management needs
5 to respond to climate impacts in order to avoid, reduce or reverse degradation. The types and intensity of
6 human land-use and climate change impacts on lands affect their carbon stocks and their ability to
7 operate as carbon sinks. In managed agricultural lands, degradation typically results in reductions of soil
8 organic carbon stocks, which also adversely affects land productivity and carbon sinks. In forest land,
9 reduction in biomass carbon stocks alone is not necessarily an indication of a reduction in carbon sinks.
10 Sustainably managed forest landscapes can have a lower biomass carbon density but the younger forests
11 can have a higher growth rate, and therefore contribute stronger carbon sinks, than older forests. Ranges
12 of carbon sinks in forest and agricultural lands are overlapping. In some cases, climate change impacts
13 may result in increased productivity and carbon stocks, at least in the short term.

14 Selective logging and thinning can maintain and enhance forest productivity and achieve co-benefits
15 when conducted with due care for the residual stand and at intensity and frequency that does not
16 exceed the rate of regrowth (Romero and Putz 2018). In contrast, unsustainable logging practices can
17 lead to stand-level degradation. For example, degradation occurs when selective logging (high-
18 grading) removes valuable large-diameter trees, leaving behind damaged, diseased, non-commercial
19 or otherwise less productive trees, reducing carbon stocks and also adversely affecting subsequent
20 forest recovery (Belair and Ducey 2018; Nyland 1992).

21 Sustainable forest management is defined using several criteria (see above) and its implementation
22 will typically involve trade-offs among these criteria. The conversion of primary forests to sustainably
23 managed forest ecosystems increases relevant economic, social and other functions but often with
24 adverse impacts on biodiversity (Barlow et al. 2007). In regions with infrequent or no stand replacing

1 natural disturbances, the timber yield per hectare harvested in managed secondary forests is typically
2 lower than the yield per hectare from the first harvest in the primary forest (Romero and Putz 2018).

3 The sustainability of timber yield has been achieved in temperate and boreal forests where
4 intensification of management has resulted in increased growing stocks and increased harvest rates in
5 countries where forests had previously been overexploited (Henttonen et al. 2017; Kauppi et al. 2018).
6 However, intensification of management to increase forest productivity can be associated with
7 reductions in biodiversity. For example, when increased productivity is achieved by periodic thinning
8 and removal of trees that would otherwise die due to competition, thinning reduces the amount of
9 dead organic matter of snags and coarse woody debris that can provide habitat and this loss reduces
10 biodiversity (Spence 2001; Ehnström 2001) and forest carbon stocks (Russell et al. 2015; Kurz et al.
11 2013). Recognition of adverse biodiversity impacts of high yield forestry is leading to modified
12 management aimed at increasing habitat availability through, for example, variable retention logging
13 and continuous cover management (Roberts et al. 2016) and through the re-introduction of fire
14 disturbances in landscapes where fires have been suppressed (Allen et al. 2002). Biodiversity losses
15 are also observed during the transition from primary to managed forests in tropical regions (Barlow et
16 al. 2007) where tree species diversity can be very high, e.g. in the Amazon region about 16,000 tree
17 species are estimated to exist (ter Steege et al. 2013).

18 Forest certification schemes have been used to document SFM outcomes (Rametsteiner and Simula
19 2003) by assessing a set of criteria and indicators (e.g., Lindenmayer et al. 2000). While many of the
20 certified forests are found in temperate and boreal countries (Rametsteiner and Simula 2003;
21 MacDicken et al. 2015), examples from the tropics also show that SFM can improve outcomes. For
22 example, selective logging emits 6% of the tropical GHG annually and improved logging practices
23 can reduce emissions by 44 % while maintaining timber production (Ellis et al. 2019). In the Congo
24 Basin, implementing reduced impact logging (RIL-C) practices can cut emissions in half without
25 reducing the timber yield (Umunay et al. 2019). SFM adoption depends on the socio-economic and
26 political context and its improvement depends mainly on better reporting and verification (Siry et al.
27 2005).

28 The successful implementation of SFM requires well established and functional governance,
29 monitoring, and enforcement mechanisms to eliminate deforestation, illegal logging, arson, and other
30 activities that are inconsistent with SFM principles (Nasi et al. 2011). Moreover, following human and
31 natural disturbances forest regrowth must be ensured through reforestation, site rehabilitation
32 activities or natural regeneration. Failure of forests to regrow following disturbances will lead to
33 unsustainable outcomes and long-term reductions in forest area, forest cover, carbon density, forest
34 productivity and land-based carbon sinks (Nasi et al. 2011).

35 Achieving all of the criteria of the definitions of SLM and SFM is an aspirational goal that will be
36 made more challenging where climate change impacts, such as biome shifts and increased
37 disturbances, are predicted to adversely affect future biodiversity and contribute to forest degradation
38 (Warren et al. 2018). Land management to enhance land sinks will involve trade-offs that need to be
39 assessed within their spatial, temporal and societal context.

40 **4.1.6 The human dimension of land degradation and forest degradation**

41 Studies of land and forest degradation are often biased towards biophysical aspects both in terms of its
42 processes, such as erosion or nutrient depletion, and its observed physical manifestations, such as
43 gullyng or low primary productivity. Land users' own perceptions and knowledge about land
44 conditions and degradation have often been neglected or ignored by both policy makers and scientists
45 (Reed et al. 2007; Forsyth 1996; Andersson et al. 2011). A growing body of work is nevertheless
46 beginning to focus on land degradation through the lens of local land users (Kessler and Stroosnijder
47 2006; Fairhead and Scoones 2005; Zimmerer 1993; Stocking et al. 2001) and the importance of local

1 and indigenous knowledge within land management is starting to be appreciated (Montanarella et al.
2 2018). Climate change impacts directly and indirectly the social reality, the land users, and the
3 ecosystem and vice versa. Land degradation can also have an impact on climate change (see Section
4 4.6).

5 The use and management of land is highly gendered and is expected to remain so for the foreseeable
6 future (Kristjanson et al. 2017). Women have often less formal access to land than men and less
7 influence over decisions about land, even if they carry out many of the land management tasks
8 (Jerneck 2018a; Elmhirst 2011; Toulmin 2009; Peters 2004; Agarwal 1997; Jerneck 2018b). Many
9 oft-cited sweeping statements about women's subordination in agriculture are difficult to substantiate,
10 yet it is clear that gender inequality persists (Doss et al. 2015). Even if women's access to land is
11 changing formally (Kumar and Quisumbing 2015), the practical outcome is often limited due to
12 several other factors related to both formal and informal institutional arrangements and values (Lavers
13 2017; Kristjanson et al. 2017; Djurfeldt et al. 2018). Women are also affected differently than men
14 when it comes to climate change, having lower adaptive capacities due to factors such as prevailing
15 land tenure frameworks, lower access to other capital assets and dominant cultural practices (Vincent
16 et al. 2014; Antwi-Agyei et al. 2015; Gabrielsson et al. 2013). This affects the options available to
17 women to respond to both land degradation and climate change. Indeed, access to land and other
18 assets (e.g., education and training) is key in shaping land-use and land management strategies (Liu et
19 al. 2018b; Lambin et al. 2001). Young people is another category that is often disadvantaged in terms
20 of access to resources and decision making power, even though they carry out much of the day-to-day
21 work (Wilson et al. 2017; Kosec et al. 2018; Naamwintome and Bagson 2013).

22 Land rights differ between places and are dependent on the political-economic and legal context
23 (Montanarella et al. 2018). This means there is no universally applicable best arrangement.
24 Agriculture in highly erosion prone regions requires site specific and long lasting soil and water
25 conservation measures, such as terraces (see 4.8.1), which may benefit from secure private land rights
26 (Tarfasa et al. 2018; Soule et al. 2000). Pastoral modes of production and community based forest
27 management systems are often dominated by communal land tenure arrangements, which may
28 conflict with agricultural/forestry modernization policies implying private property rights (Antwi-
29 Agyei et al. 2015; Benjaminsen and Lund 2003; Itkonen 2016; Owour et al. 2011; Gebara 2018)

30 Cultural ecosystem services, defined as the non-material benefits people obtain from ecosystems
31 through spiritual enrichment, cognitive development, reflection, recreation and aesthetic experiences
32 (Millennium Assessment 2005) are closely linked to land and ecosystems, although often
33 underrepresented in the literature on ecosystem services (Tengberg et al. 2012; Hernández-Morcillo
34 et al. 2013). Climate change interacting with land conditions can impact cultural aspects, such as
35 sense of place and sense of belonging (Olsson et al. 2014).

36 **4.2 Land degradation in the context of climate change**

37 Land degradation results from a complex chain of causes making the clear distinction between direct
38 and indirect drivers difficult. In the context of climate change, an additional complex aspect is brought
39 by the reciprocal effects that both processes have on each other (i.e. climate change influencing land
40 degradation and vice versa). In this chapter, we use the terms processes and drivers with the following
41 meanings:

42 **Processes of land degradation** are those direct mechanisms by which land is degraded and are
43 similar to the notion of "direct drivers" in the Millennium Ecosystem Assessment (MA, Millennium
44 Ecosystem Assessment, 2005) framework. In this report, a comprehensive list of land degradation
45 processes is presented in Table 4.1.

1 **Drivers of land degradation** are those indirect conditions which may drive processes of land
2 degradation and are similar to the notion of “indirect drivers” in the MA framework. Examples of
3 indirect drivers of land degradation are changes in land tenure or cash crop prices, which can trigger
4 land-use or management shifts that affect land degradation.

5 An exact demarcation between processes and drivers is not possible. Drought and fires are described
6 as drivers of land degradation in the next section but they can also be a process: for example, if
7 repeated fires deplete seed sources they can affect regeneration and succession of forest ecosystems.
8 The responses to land degradation follow the logic of the Land Degradation Neutrality concept:
9 avoiding, reducing and reversing land degradation (Orr et al. 2017b; Cowie et al. 2018).

10 In research on land degradation, climate and climate variability are often intrinsic factors. The role of
11 climate change, however, is less articulated. Depending on what conceptual framework is used,
12 climate change is understood either as a process or a driver of land degradation, and sometimes both.

13 **4.2.1 Processes of land degradation**

14 A large array of interactive physical, chemical, biological and human processes led to what we define
15 in this report as land degradation (Johnson and Lewis 2007). The biological productivity, ecological
16 integrity (which encompasses both functional and structural attributes of ecosystems) or the human
17 value (which includes any benefit that people get from the land) of a given territory can deteriorate as
18 the result of processes triggered at scales that range from a single furrow (e.g., water erosion under
19 cultivation) to the landscape level (e.g., salinisation through raising groundwater levels under
20 irrigation). While pressures leading to land degradation are often exerted on specific components of
21 the land systems (i.e., soils, water, biota), once degradation processes start, other components become
22 affected through cascading and interactive effects. For example, different pressures and degradation
23 processes can have convergent effects, as can be the case of overgrazing leading to wind erosion,
24 landscape drainage resulting in wetland drying, and warming causing more frequent burning; all of
25 which can independently lead to reductions of the soil organic matter pools as second order process.
26 Still, the reduction of organic matter pools is also a first order process triggered directly by the effects
27 of rising temperatures (Crowther et al., 2016) as well as other climate changes such as precipitation
28 shifts (Viscarra Rossel et al. 2014). Beyond this complexity, a practical assessment of the major land
29 degradation processes helps to reveal and categorise the multiple pathways in which climate change
30 exerts a degradation pressure (Table 4.1).

31 Conversion of freshwater wetlands to agricultural land has historically been a common way of
32 increasing the area of arable land. Despite the small areal extent (~1% of the earth’s surface (Hu et al.
33 2017; Dixon et al. 2016)), freshwater wetlands provide a very large number of ecosystem services,
34 such as groundwater replenishment, flood protection, and nutrient retention, and are biodiversity
35 hotspots (Reis et al. 2017; Darrah et al. 2019; Montanarella et al. 2018). The loss of wetlands since
36 1900 has been estimated at ~55% globally (Davidson 2014) (*low confidence*) and 35% since 1970
37 (Darrah et al. 2019) (*medium confidence*) which in many situations pose a problem for adaptation to
38 climate change. Drainage causes loss of wetlands, which can be further exacerbated by climate
39 change, and reduces the capacity to adapt to climate change (Barnett et al. 2015; Colloff et al. 2016;
40 Finlayson et al. 2017) (*high confidence*).

41 **4.2.1.1 Types of land degradation processes**

42 Land degradation processes can affect the soil, water or biotic components of the land or in their
43 respective interfaces (Table 4.1). Across land degradation processes, those affecting the soil have
44 received more attention. The most widespread and studied land degradation processes affecting soils
45 are water and wind erosion, which have accompanied agriculture since its onset and are still dominant
46 (Table 4.1). Degradation through erosion processes is not restricted to soil loss in detachment areas
47 but includes impacts on transport and deposition areas as well (less commonly, deposition areas can

1 have their soils improved by these inputs). Larger scale degradation processes related to the whole
2 continuum of soil erosion, transport and deposition include dune field expansion/displacement,
3 development of gully networks and the accumulation of sediments (siltation) of natural and artificial
4 water bodies (Poesen and Hooke 1997; Ravi et al. 2010). Long-distance sediment transport during
5 erosion events can have remote effects on land systems as documented for the fertilisation effect of
6 African dust on the Amazon (Yu et al. 2015).

7 Coastal erosion represents a special case among erosional, with reports linking it to climate change.
8 While human interventions in coastal areas (e.g., expansion of shrimp farms) and rivers (e.g.,
9 upstream dams cutting coastal sediment supply), and economic activities causing land subsidence
10 (Keogh and Törnqvist 2019; Allison et al. 2016) are dominant human drivers, storms and sea level
11 rise have already left a significant global imprint on coastal erosion (Mentaschi et al. 2018). Recent
12 projections that take into account geomorphological and socioecological feedbacks suggest that
13 coastal wetlands may not get reduced by sea level rise if their inland growth is accommodated with
14 proper management actions (Schuerch et al. 2018a).

15 Other physical degradation process in which no material detachment and transport are involved
16 include soil compaction, hardening, sealing and any other mechanism leading to the loss of porous
17 space crucial for holding and exchanging air and water (Hamza and Anderson 2005). A very extreme
18 case of degradation through pore volume loss, manifested at landscape or larger scales, is ground
19 subsidence. Typically caused by the lowering of groundwater or oil levels, subsidence involves a
20 sustained collapse of the ground surface, which can lead to other degradation processes such as
21 salinisation and permanent flooding. Chemical soil degradation processes include relatively simple
22 changes, like nutrient depletion resulting from the imbalance of nutrient extraction on harvested
23 products and fertilisation, and more complex ones, such as acidification and increasing metal toxicity.
24 Acidification in croplands is increasingly driven by excessive nitrogen fertilisation and to a lower
25 extent by the depletion of cation like calcium, potassium or magnesium through exports in harvested
26 biomass (Guo et al. 2010). One of the most relevant chemical degradation processes of soils in the
27 context of climate change is the depletion of its organic matter pool. Reduced in agricultural soils
28 through the increase of respiration rates by tillage and the decline of belowground plant biomass
29 inputs, soil organic matter pools have been diminished also by the direct effects of warming, not only
30 in cultivated land but also under natural vegetation (Bond-Lamberty et al. 2018). Debate persists,
31 however, on whether in more humid and carbon rich ecosystems the simultaneous stimulation of
32 decomposition and productivity may result in the lack of effects on soil carbon (Crowther et al. 2016;
33 van Gestel et al. 2018). In the case of forests, harvesting, particularly if it is exhaustive as in the case
34 of the use of residues for energy generation, can also lead to organic matter declines (Achat et al.
35 2015). Affected by many other degradation processes (e.g. wildfire increase, salinisation) and having
36 negative effects on other pathways of soil degradation (e.g. reduced nutrient availability, metal
37 toxicity). Soil organic matter can be considered a “hub” of degradation processes and a critical link
38 with the climate system (Minasny et al. 2017).

39 Land degradation processes can also start from alterations in the hydrological system that are
40 particularly important in the context of climate change. Salinisation, although perceived and reported
41 in soils, is typically triggered by water table-level rises driving salts to the surface under dry to sub-
42 humid climates (Schofield and Kirkby 2003). While salty soils occur naturally under these climates
43 (primary salinity), human interventions have expanded their distribution (secondary salinity with
44 irrigation without proper drainage being the predominant cause of salinisation (Rengasamy 2006).
45 Yet, it has also taken place under non-irrigated conditions where vegetation changes (particularly dry
46 forest clearing and cultivation) had reduced the magnitude and depth of soil water uptake, triggering
47 water table rises towards the surface. Changes in evapotranspiration and rainfall regimes can
48 exacerbate this process (Schofield and Kirkby 2003). Salinisation can also result from the intrusion of

1 sea water into coastal areas both as a result of sea level rise and ground subsidence (Colombani et al.
2 2016).

3 Recurring flood and waterlogging episodes (Bradshaw et al. 2007; Poff 2002), and the more chronic
4 expansion of wetlands over dryland ecosystems are mediated by the hydrological system, on
5 occasions aided by geomorphological shifts as well (Kirwan et al. 2011). This is also the case for the
6 drying of continental water bodies and wetlands, including the salinisation and drying of lakes and
7 inland seas (Anderson et al. 2003; Micklin 2010; Herbert et al. 2015). In the context of climate
8 change, the degradation of peatland ecosystems is particularly relevant given their very high carbon
9 storage and their sensitivity to changes in soils, hydrology and/or vegetation (Leifeld and Menichetti
10 2018). Drainage for land-use conversion together with peat mining are major drivers of peatland
11 degradation, yet other factors such as the extractive use of their natural vegetation and the interactive
12 effects of water table levels and fires (both sensitive to climate change) are important (Hergoualc'h et
13 al. 2017a; Lilleskov et al. 2019).

14 The biotic components of the land can also be the focus of degradation processes. Vegetation clearing
15 processes associated with land-use changes are not limited to deforestation but include other natural
16 and seminatural ecosystems such as grasslands (the most cultivated biome on Earth), as well as dry
17 steppes and shrublands, which give place to croplands, pastures, urbanisation or just barren land. This
18 clearing process is associated with net C losses from the vegetation and soil pool. Not all biotic
19 degradation processes involve biomass losses. Woody encroachment of open savannahs involve the
20 expansion of woody plant cover and/or density over herbaceous areas and often limits the secondary
21 productivity of rangelands (Asner et al. 2004, Anadon et al. 2014). These processes have been
22 accelerated since the mid-1800s over most continents (Van Auken 2009). Change in plant
23 composition of natural or semi-natural ecosystems without any significant vegetation structural
24 changes is another pathway of degradation affecting rangelands and forests. In rangelands, selective
25 grazing and its interaction with climate variability and/or fire can push ecosystems to new
26 compositions with lower forage value and higher proportion of invasive species (Illius and O'Connor
27 1999, Sasaki et al. 2007), in some cases with higher carbon sequestration potential, yet with very
28 complex interactions between vegetation and soil carbon shifts (Piñeiro et al. 2010). In forests,
29 extractive logging can be a pervasive cause of degradation leading to long-term impoverishment and
30 in extreme cases, a full loss of the forest cover through its interaction with other agents such as fires
31 (Foley et al. 2007) or progressive intensification of land use. Invasive alien species are another source
32 of biological degradation. Their arrival into cultivated systems is constantly reshaping crop production
33 strategies making agriculture unviable on occasions. In natural and seminatural systems such as
34 rangelands, invasive plant species not only threaten livestock production through diminished forage
35 quality, poisoning and other deleterious effects, but have cascading effects on other processes such as
36 altered fire regimes and water cycling (Brooks et al. 2004). In forests, invasions affect primary
37 productivity and nutrient availability, change fire regimes, and alter species composition, resulting in
38 long term impacts on carbon pools and fluxes (Peltzer et al. 2010).

39 Other biotic components of ecosystems have been shown as a focus of degradation processes.
40 Invertebrate invasions in continental waters can exacerbate other degradation processes such as
41 eutrophication, which is the over enrichment of nutrients leading to excessive algal growth (Walsh et
42 al. 2016a). Shifts in soil microbial and mesofaunal composition, which can be caused by pollution
43 with pesticides or nitrogen deposition but also by vegetation or disturbance regime shifts, alter many
44 soil functions including respiration rates and C release to the atmosphere (Hussain et al. 2009;
45 Crowther et al. 2015). The role of the soil biota modulating the effects of climate change on soil
46 carbon have been recently demonstrated (Ratcliffe et al. 2017), highlighting the importance of this
47 less known component of the biota as a focal point of land degradation. Of special relevance as both
48 indicators and agents of land degradation recovery are mycorrhiza, which are root associated fungal

1 organisms (Asmelash et al. 2016; Vasconcellos et al. 2016). In natural dry ecosystems, biological soil
2 crusts composed by a broad range of organisms including mosses are a particularly sensitive focus for
3 degradation (Field et al. 2010) with evidenced sensitivity to climate change (Reed et al. 2012).

4 **4.2.1.2 Land degradation processes and climate change**

5 While the subdivision of individual processes is challenged by their strong interconnectedness, it
6 provides a useful setting to identify the most important “focal points” of climate change pressures on
7 land degradation. Among land degradation processes those responding more directly to climate
8 change pressures include all types of erosion and soil organic matter declines (soil focus), salinisation,
9 sodification and permafrost thawing (soil/water focus), waterlogging of dry ecosystems and drying of
10 wet ecosystems (water focus), and a broad group of biological mediated processes like woody
11 encroachment, biological invasions, pest outbreaks (biotic focus), together with biological soil crust
12 destruction and increased burning (soil/biota focus) (Table 4.1). Processes like ground subsidence can
13 be affected by climate change indirectly through sea level rise (Keogh and Törnqvist 2019).

14 Even when climate change exerts a direct pressure on degradation processes, it can be a secondary
15 driver subordinated to other overwhelming human pressures. Important exceptions are three processes
16 in which climate change is a dominant global or regional pressure and the main driver of their current
17 acceleration. These are coastal erosion as affected by sea level rise and increased storm
18 frequency/intensity (*high agreement, medium evidence*) (Johnson et al. 2015; Alongi 2015; Harley et
19 al. 2017a; Nicholls et al. 2016), permafrost thawing responding to warming (*high agreement, robust
20 evidence*) (Liljedahl et al. 2016; Peng et al. 2016; Batir et al. 2017) and increased burning responding
21 to warming and altered precipitation regimes (*high agreement, robust evidence*) (Jolly et al. 2015;
22 Abatzoglou and Williams 2016; Taufik et al. 2017; Knorr et al. 2016). The previous assessment
23 highlights the fact that climate change not only exacerbates many of the well acknowledged ongoing
24 land degradation processes of managed ecosystems (i.e., croplands and pastures), but becomes a
25 dominant pressure that introduces novel degradation pathways in natural and seminatural ecosystems.
26 Climate change has influenced species invasions and the degradation that they cause by enhancing the
27 transport, colonisation, establishment, and ecological impact of the invasive species, but also by
28 impairing their control practices (*medium agreement, medium evidence*) (Hellmann et al. 2008).

1 **Table 4.1 Major land degradation processes and their connections with climate change.** For each process a “focal point” (soil, water, biota) on which degradation
 2 occurs first place is indicated, acknowledging that most processes propagate to other land components and cascade into or interact with some of the other processes
 3 listed below. The impact of climate change on each process is categorised based on the proximity (very direct = high, very indirect=low) and dominance
 4 (dominant=high, subordinate to other pressures = low) of effects. The major effects of climate change on each process are highlighted together with the
 5 predominant pressures from other drivers. Feedbacks of land degradation processes on climate change are categorized according to the intensity (very
 6 intense=high, subtle=low) of the chemical (greenhouse gases emissions or capture) or physical (energy and momentum exchange, aerosol emissions) effects.
 7 Warming effects are indicated in red and cooling effects in blue. Specific feedbacks on climate change are highlighted.

| Processes | Focal point | Impacts of Climate Change | | Feedbacks on Climate Change | | | | | |
|-----------------|-------------|---------------------------|-----------|---|--|---------------|------------------|------|---|
| | | proximity | dominance | intensity of chemical effects | intensity of physical effects | global extent | Specific Impacts | | |
| | | | | Climate Change pressures | Other pressures | | | | |
| Wind erosion | Soil | high | medium | Altered wind/drought patterns (<i>high confidence</i> on effect, <i>medium-low confidence</i> on trend) (1). Indirect effect through vegetation type and biomass production shifts | Tillage, leaving low cover, overgrazing, deforestation/vegetation clearing, large plot sizes, vegetation and fire regime shifts | low | medium | high | Radiative cooling by dust release (<i>medium confidence</i>). Ocean and land fertilisation and C burial (<i>medium confidence</i>). Albedo increase. Dust effect as condensation nuclei (19). |
| Water erosion | Soil | high | medium | Increasing rainfall intensity (<i>high confidence</i> on effect and trend) (2). Indirect effects on fire frequency/intensity, permafrost thawing, biomass production. | Tillage, cultivation leaving low cover, overgrazing, deforestation/vegetation clearing, vegetation burning, poorly designed roads and paths. | medium | medium | high | Net C release. Net release is probably less than site-specific loss due to deposition and burial (<i>high confidence</i>). Albedo increase (20). |
| Coastal erosion | Soil/Water | high | high | Sea level rise, increasing intensity/frequency of storm surges (<i>high confidence</i> on effects and trends)(3) | Retention of sediments by upstream dams, Coastal aquaculture, Elimination of mangrove forests, Subsidence | high | low | low | Release of old buried C pools (<i>medium confidence</i>)(21). |

| | | | | | | | | | |
|---------------------------------|------------|------|--------|---|---|----------|-----|--------|--|
| Subsidence | Soil/Water | low | low | Indirect through increasing drought leading to higher ground water use. Indirect through enhanced decomposition (e.g. through drainage) in organic soils. | Groundwater depletion / overpumping. Peatland drainage. | low/high | low | low | Unimportant in the case of groundwater depletion. Very high net C release in the case of drained peatlands |
| Compaction/Hardening | Soil | low | low | Indirect through reduced organic matter content. | Land use conversion, machinery overuse, intensive grazing, poor tillage/grazing management (e.g. under wet or waterlogged conditions) | low | low | medium | Contradictory effects of reduced aeration on N ₂ O emissions |
| Nutrient depletion | Soil | low | low | Indirect (e.g. shifts in cropland distribution, BECCS) | Insufficient replenishment of harvested nutrients | low | low | medium | Net C release due shrinking SOC pools. Larger reliance on soil liming with associated CO ₂ releases. |
| Acidification/Overfertilisation | Soil | low | low | Indirect (e.g. shifts in cropland distribution, BECCS). Sulfidic wetland drying due to increased drought as special direct effect. | High N fertilisation. High cation depletion. Acid rain/deposition | medium | low | medium | N ₂ O release from overfertilised soils, increased by acidification. Inorganic C release from acidifying soils (<i>medium to high confidence</i>) (22). |
| Pollution | Soil/Biota | low | low | Indirect (e.g. increased pest and weed incidence) | Intensifying chemical control of weed and pests | low | low | medium | Unknown, probably unimportant. |
| Organic matter decline | Soil | high | medium | Warming accelerates soil respiration rates (<i>medium confidence</i> on effects and trends) (4). Indirect effects through changing quality of plant litter or fire/waterlogging regimes. | Tillage. reduced plant input to soil. Drainage of waterlogged soils. Influenced by most of the other soil degradation processes. | high | low | high | Net C release (<i>high confidence</i>)(23). |
| Metal toxicity | Soil | low | low | Indirect | High cation depletion, fertilisation, mining activities | low | low | low | unknown, probably unimportant. |

| | | | | | | | | | |
|--|--------------|------|--------|--|---|--------|--------|--------|---|
| Salinisation | Soil / Water | High | low | Sea level rise (<i>high confidence</i> on effects and trends) (5). Water balance shifts (<i>medium confidence</i> on effects and trends) (6). Indirect effects through irrigation expansion. | Irrigation without good drainage infrastructure. Deforestation and water table level raises under dryland agriculture | low | medium | medium | Reduced methane emissions with high sulfate load. Albedo increase. |
| Sodification (increased sodium and associated physical degradation in soils) | Soil / Water | High | low | Water balance shifts (<i>medium confidence</i> on effects and trends) (7). Indirect effects through irrigation expansion. | Poor water management | low | medium | low | Net C release due to soil structure and organic matter dispersion. Albedo increase. |
| Permafrost thawing | Soil / Water | High | high | Warming (<i>very high confidence</i> on effects and trends) (8), seasonality shifts and accelerated snow melt leading to higher erosivity. | | high | low | high | Net C release. CH ₄ release (<i>high confidence</i>)(24). |
| Waterlogging of dry systems | Water | High | medium | Water balance shifts (<i>medium confidence</i> on effects and trends) (9). Indirect effects through vegetation shifts. | Deforestation. Irrigation without good drainage infrastructure | medium | medium | low | CH ₄ release. Albedo decrease |
| Drying of continental waters/wetland/lowlands | Water | High | medium | Increasing extent and duration of drought (<i>high confidence</i> on effects, <i>medium confidence</i> on trends) (10). Indirect effects through vegetation shifts. | Upstream surface and groundwater water consumption. Intentional drainage. Trampling/overgrazing. | medium | medium | medium | Net C release. N ₂ O release. Albedo increase |
| Flooding | Water | High | medium | Sea level raise, increasing intensity/frequency of storm surges, increasing rainfall intensity causing flash floods (<i>high confidence</i> on effects and trends)(11). | Land clearing. Increasing impervious surface. Transport infrastructure. | medium | medium | low | CH ₄ and N ₂ O release. Albedo decrease |
| Eutrophication of continental waters | Water/Biota | Low | low | Indirect through warming effects on N losses from the land or climate change effects on erosion rates. Interactive effects of warming and nutrient loads on algal blooms. | Excess fertilisation. Erosion. Poor management of livestock/human sewage. | medium | low | low | CH ₄ and N ₂ O release. |

| | | | | | | | | | |
|--------------------------------------|------------|------|--------|--|---|--------|------|--------|--|
| Woody encroachment | Biota | High | medium | Rainfall shifts (<i>medium confidence</i> on effects and trends), CO ₂ rise (<i>medium confidence</i> on effects, <i>very high confidence</i> on trends)(12). | Overgrazing. Altered fire regimes, fire suppression. Invasive alien species. | high | high | high | Net C storage. Albedo decrease |
| Species loss, compositional shifts | Biota | High | medium | Habitat loss as a result of climate shifts (<i>medium confidence</i> on effects and trends) (13). | Selective grazing and logging causing plant species loss, Pesticides causing soil microbial and soil faunal losses, Large animal extinctions, Interruption of disturbance regimes | low | low | medium | Unknown. |
| Soil microbial and mesofaunal shifts | Biota | High | low | Habitat loss as a result of climate shifts (<i>medium confidence</i> on effects and trends) (14). | Altered fire regimes, nitrogen deposition, pesticide pollution, vegetation shifts, disturbance regime shifts | low | low | medium | Unknown. |
| Biological soil crust destruction | Biota/Soil | High | medium | Warming. Changing rainfall regimes. (<i>medium confidence</i> on effects, high confidence and trends). Indirect through fire regime shifts and/or invasions (15). | Overgrazing and trampling. Land use conversion. | low | high | high | Radiative cooling through albedo rise and dust release (<i>high confidence</i>)(25). |
| Invasions | Biota | High | medium | Habitat gain as a result of climate shifts (<i>medium confidence</i> on effects and trends) (16). | Intentional and unintentional species introductions. | low | low | medium | Unknown. |
| Pest outbreaks | Biota | High | medium | Habitat gain and accelerated reproduction as a result of climate shifts (<i>medium confidence</i> on effects and trends) (17). | Large scale monocultures. Poor pest management practices. | medium | low | medium | Net C release. |

| | | | | | | | | | |
|-------------------|------------|------|------|---|---|------|--------|--------|--|
| Increased burning | Soil/Biota | High | high | Warming, drought, shifting precipitation regimes, also wet spells rising fuel load. (<i>high confidence</i> on effects and trends) (18). | Fire suppression policies increasing wildfire intensity. Increasing use of fire for rangeland management. Agriculture introducing fires in humid climates without previous fire history. Invasions. | high | medium | medium | Net C release. CO, CH ₄ , N ₂ O release. Albedo increase. (<i>high confidence</i>). Long term decline of NPP in non-adapted ecosystems (26). |
|-------------------|------------|------|------|---|---|------|--------|--------|--|

1

2 References in table 4.1:

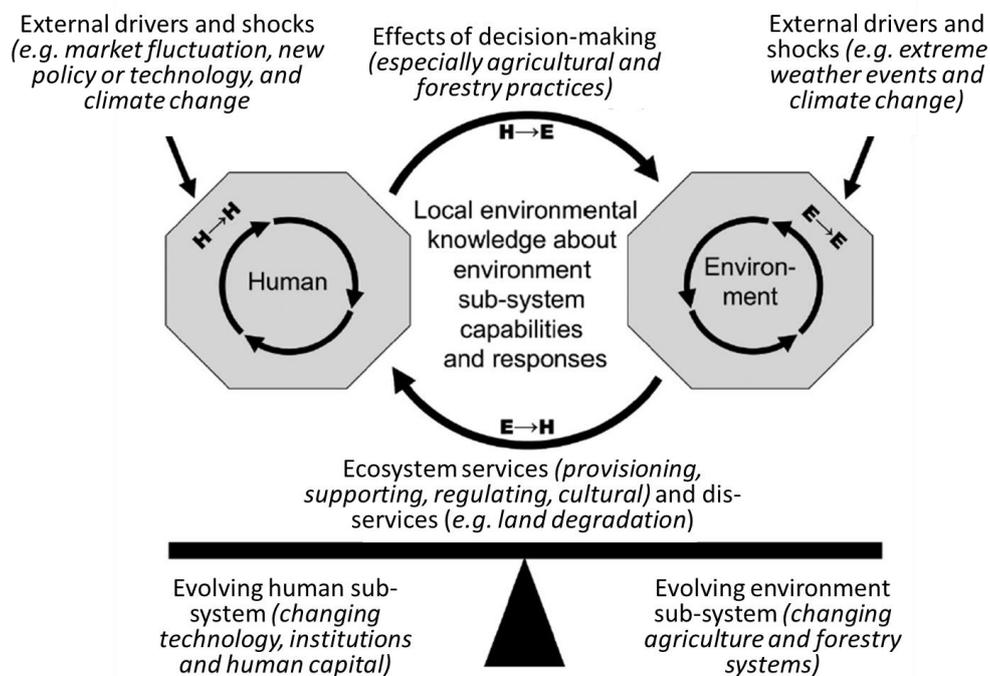
3 (1) (Barring et al. 2003; Munson et al. 2011; Sheffield et al. 2012), (2) (Nearing et al. 2004b; Shakesby 2011; Panthou et al. 2014), (3) (Johnson et al. 2015; Alongi 2015;
 4 Harley et al. 2017b), (4) (Bond-Lamberty et al. 2018; Crowther et al. 2016; van Gestel et al. 2018), (5) (Colombani et al. 2016), (6) (Schofield and Kirkby 2003; Aragüés et
 5 al. 2015; Benini et al. 2016), (7) (Jobbágy et al. 2017), (8) (Liljedahl et al. 2016; Peng et al. 2016; Batir et al. 2017), (9) (Piovano et al. 2004; Osland et al. 2016), (10)
 6 (Burkett and Kusler 2000; Nielsen and Brock 2009; Johnson et al. 2015; Green et al. 2017), (11) (Panthou et al. 2014; Arnell and Gosling 2016; Vitousek et al. 2017), (12)
 7 (Van Auken 2009; Wigley et al. 2010), (13) (Vincent et al. 2014; Gonzalez et al. 2010; Scheffers et al. 2016), (14) (Pritchard 2011; Ratcliffe et al. 2017), (15) (Reed et al.
 8 2012; Maestre et al. 2013), (16) (Hellmann et al. 2008; Hulme 2017), (17) (Pureswaran et al. 2015; Cilas et al. 2016; Macfadyen et al. 2018), (18) (Jolly et al. 2015;
 9 Abatzoglou and Williams 2016; Taufik et al. 2017; Knorr et al. 2016), (19) (Davin et al. 2010; Pinty et al. 2011), (20) (Wang et al. 2017b; Chappell et al. 2016), (21)
 10 (Pendleton et al. 2012), (22) (Oertel et al. 2016), (23) (Houghton et al. 2012; Eglin et al. 2010), (24) (Schuur et al. 2015; Christensen et al. 2004; Walter Anthony et al. 2016;
 11 Abbott et al. 2016), (25) (Belnap, Walker, Munson, & Gill, 2014; Rutherford et al., 2017), (26) (Page et al. 2002; Pellegrini et al. 2018)

12

13

1 4.2.2 Drivers of land degradation

2 Drivers of land degradation and land improvement are many and they interact in multiple ways.
 3 Figure 4.2, illustrates how some of the most important drivers interact with the land users. It is
 4 important to keep in mind that both natural and human factors can drive both degradation and
 5 improvement (Kiage 2013; Bisaro et al. 2014).



6

7 **Figure 4.2 Schematic representation of the interactions between the human and environmental**
 8 **components of the land system showing decision making and ecosystem services as the key linkages**
 9 **between the components (moderated by an effective system of local and scientific knowledge), and**
 10 **indicating how the rates of change and the way these linkages operate must be kept broadly in balance for**
 11 **functional coevolution of the components. Modified with permission from (Stafford Smith et al. 2007).**

12 Land degradation is driven by the entire spectrum of factors, from very short and intensive events
 13 such as individual rain storms of 10 minutes removing topsoil or initiating a gully or a landslide
 14 (Coppus and Imeson 2002; Morgan 2005b) to century scale slow depletion of nutrients or loss of soil
 15 particles (Johnson and Lewis 2007, p. 5-6). But instead of focusing on absolute temporal variations,
 16 the drivers of land degradation can be assessed in relation to the rates of possible recovery.
 17 Unfortunately, this is impractical to do in a spatially explicit way because rates of soil formation is
 18 difficult to measure due the slow rate, usually < 5mm/century (Delgado and Gómez 2016). Studies
 19 suggest that erosion rates of conventionally tilled agricultural fields exceed the rate at which soil is
 20 generated by one to two orders of magnitude (Montgomery 2007a).

21 The landscape effects of gully erosion from one short intensive rainstorm can persist for decades and
 22 centuries (Showers 2005). Intensive agriculture under the Roman Empire in occupied territories in
 23 France is still leaving its marks and can be considered an example of irreversible land degradation
 24 (Dupouey et al. 2002).

25 The climate change related drivers of land degradation are both gradual changes of temperature,
 26 precipitation, and wind as well as changes of the distribution and intensity of extreme events (Lin et
 27 al. 2017). Importantly, these drivers can act in two directions: land improvement and land

1 degradation. Increasing CO₂ levels in the atmosphere is a driver of land improvement even if the net
2 effect is modulated by other factors, such as the availability of nitrogen (Terrer et al. 2016) and water
3 (Gerten et al. 2014; Settele et al. 2015; Girardin et al. 2016).

4 The gradual and planetary changes that can cause land degradation/improvement have been studied by
5 global integrated models and Earth observation technologies. Studies of global land suitability for
6 agriculture suggest that climate change will increase the area suitable for agriculture by 2100 in the
7 Northern high latitudes by 16% (Ramankutty et al. 2002) or 5.6 million km² (Zabel et al. 2014), while
8 tropical regions will experience a loss (Ramankutty et al. 2002; Zabel et al. 2014).

9 Temporal and spatial patterns of tree mortality can be used as an indicator of climate change impacts
10 on terrestrial ecosystems. Episodic mortality of trees occur naturally even without climate change, but
11 more widespread spatio-temporal anomalies can be a sign of climate induced degradation (Allen et al.
12 2010). In the absence of systematic data on tree mortality, a comprehensive meta-analysis of 150
13 published articles suggests that increasing tree mortality around the world can be attributed to
14 increasing drought and heat stress in forests worldwide (Allen et al. 2010).

15 Other and more indirect drivers can be a wide range of factors such as demographic changes,
16 technological change, changes of consumption patterns and dietary preferences, political and
17 economic changes, and social changes (Mirzabaev et al. 2016). It is important to stress that there are
18 no simple or direct relationships between underlying drivers and land degradation, such as poverty or
19 high population density, that are necessarily causing land degradation (Lambin et al. 2001). However,
20 drivers of land degradation need to be studied in the context of spatial, temporal, economic,
21 environmental and cultural aspects (Warren 2002). Some analyzes suggest an overall negative
22 correlation between population density and land degradation (Bai et al. 2008) but we find many local
23 examples of both positive and negative relationships (Brandt et al. 2018a, 2017). Even if there are
24 correlations in one or the other direction, causality is not always the same.

25 Land degradation is inextricably linked to several climate variables, such as temperature,
26 precipitation, wind, and seasonality. This means that there are many ways in which climate change
27 and land degradation are linked. The linkages are better described as a web of causality than a set of
28 cause – effect relationships.

29 **4.2.3 Attribution in the case of land degradation**

30 The question here is whether or not climate change can be attributed to land degradation and vice
31 versa. Land degradation is a complex phenomenon often affected by multiple factors such as climatic
32 (rainfall, temperature, and wind), abiotic ecological factors (e.g. soil characteristics and topography),
33 type of land use (e.g. farming of various kinds, forestry, or protected area), and land management
34 practices (e.g. tilling, crop rotation, and logging/thinning). Therefore, attribution of land degradation
35 to climate change is extremely challenging. Because land degradation is highly dependent on land
36 management, it is even possible that climate impacts would trigger land management changes
37 reducing or reversing land degradation, sometimes called transformational adaptation (Kates et al.
38 2012). There is not much research on attributing land degradation explicitly to climate change, but
39 there is more on climate change as a threat multiplier for land degradation. However, it is in some
40 cases possible to infer climate change impacts on land degradation both theoretically and empirically.
41 Section 4.2.3.1 will outline the potential direct linkages of climate change on land degradation based
42 on current theoretical understanding of land degradation processes and drivers. Section 4.2.3.2 will
43 investigate possible indirect impacts on land degradation.

44 **4.2.3.1 Direct linkages with climate change**

45 The most important direct impacts of climate change on land degradation are the results of increasing
46 temperatures, changing rainfall patterns, and intensification of rainfall. These changes will in various

1 combinations cause changes in erosion rates and the processes driving both increases and decreases of
2 soil erosion. From an attribution point of view, it is important to note that projections of precipitation
3 are in general more uncertain than projections of temperature changes (Murphy et al. 2004; Fischer
4 and Knutti 2015; IPCC 2013a). Precipitation involves local processes of larger complexity than
5 temperature and projections are usually less robust than those for temperature (Giorgi and Lionello
6 2008; Pendergrass 2018).

7 Theoretically the intensification of the hydrological cycle as a result of human induced climate change
8 is well established (Guerreiro et al. 2018; Trenberth 1999; Pendergrass et al. 2017; Pendergrass and
9 Knutti 2018) and also empirically observed (Blenkinsop et al. 2018; Burt et al. 2016a; Liu et al. 2009;
10 Bindoff et al. 2013). AR5 WGI concluded that heavy precipitation events have increased in
11 frequency, intensity, and/or amount since 1950 (*likely*) and that further changes in this direction are
12 *likely to very likely* during the 21st century (IPCC 2013). The IPCC Special Report on 1.5°C
13 concluded that human-induced global warming has already caused an increase in the frequency,
14 intensity and/or amount of heavy precipitation events at the global scale (Hoegh-Guldberg et al.
15 2018). As an example, in central India, there has been a threefold increase in widespread extreme rain
16 events during 1950–2015 which has influenced several land degradation processes, not least soil
17 erosion (Burt et al. 2016b). In Europe and North America, where observation networks are dense and
18 having long time series, it is *likely* that the frequency or intensity of heavy rainfall have increased
19 (IPCC 2013b). It is also expected that seasonal shifts and cycles such as monsoons and ENSO (see
20 Glossary) will further increase the intensity of rainfall events (IPCC 2013).

21 When rainfall regimes change, it is expected to drive changes in vegetation cover and composition,
22 which may be a cause of land degradation in and of itself, as well as impacting other aspects of land
23 degradation. Vegetation cover, for example is a key factor in determining soil loss through both water
24 (Nearing et al. 2005) and wind erosion (Shao 2008). Changing rainfall regimes also affect below-
25 ground biological processes, such as fungi and bacteria (Meisner et al. 2018; Shuab et al. 2017;
26 Asmelash et al. 2016).

27 Changing snow accumulation and snow melt alter both volume and timing of hydrological flows in
28 and from mountain areas (Brahney et al. 2017; Lutz et al. 2014), with potentially large impacts on
29 downstream areas. Soil processes are also affected by changing snow conditions by affecting the
30 partitioning between evaporation and streamflow and between subsurface flow and surface runoff
31 (Barnhart et al. 2016). Rainfall intensity is a key climatic driver of soil erosion. Early modelling
32 studies and theory suggest that light rainfall events will decrease while heavy rainfall events increase
33 at about 7% per degree of warming (Liu et al. 2009; Trenberth 2011). Such changes result in increases
34 in the intensity of rainfall which increase the erosive power of rainfall (erosivity) and hence increase
35 the likelihood of water erosion. Increases in rainfall intensity can even exceed the rate of increase of
36 atmospheric moisture content (Liu et al. 2009; Trenberth 2011). Erosivity is highly correlated to the
37 product of total rainstorm energy and the maximum 30 minute rainfall intensity of the storm (Nearing
38 et al. 2004a) and increases of erosivity will exacerbate water erosion substantially (Nearing et al.
39 2004a). However, the effects will not be uniform but highly variable across regions (Almagro et al.
40 2017; Mondal et al. 2016). Several empirical studies around the world have shown the increasing
41 intensity of rainfall (IPCC 2013b; Ma et al. 2015, 2017) and also suggest that this will be accentuated
42 with future increasing warming (Cheng and AghaKouchak 2015; Burt et al. 2016b; O’Gorman 2015).

43 The very comprehensive database of direct measurements of water erosion presented by García-Ruiz
44 et al. (2015) contains 4377 entries (North America: 2776, Europe: 847, Asia: 259, Latin America:
45 237, Africa: 189, Australia & Pacific: 67), even though not all entries are complete (Figure 4.3).

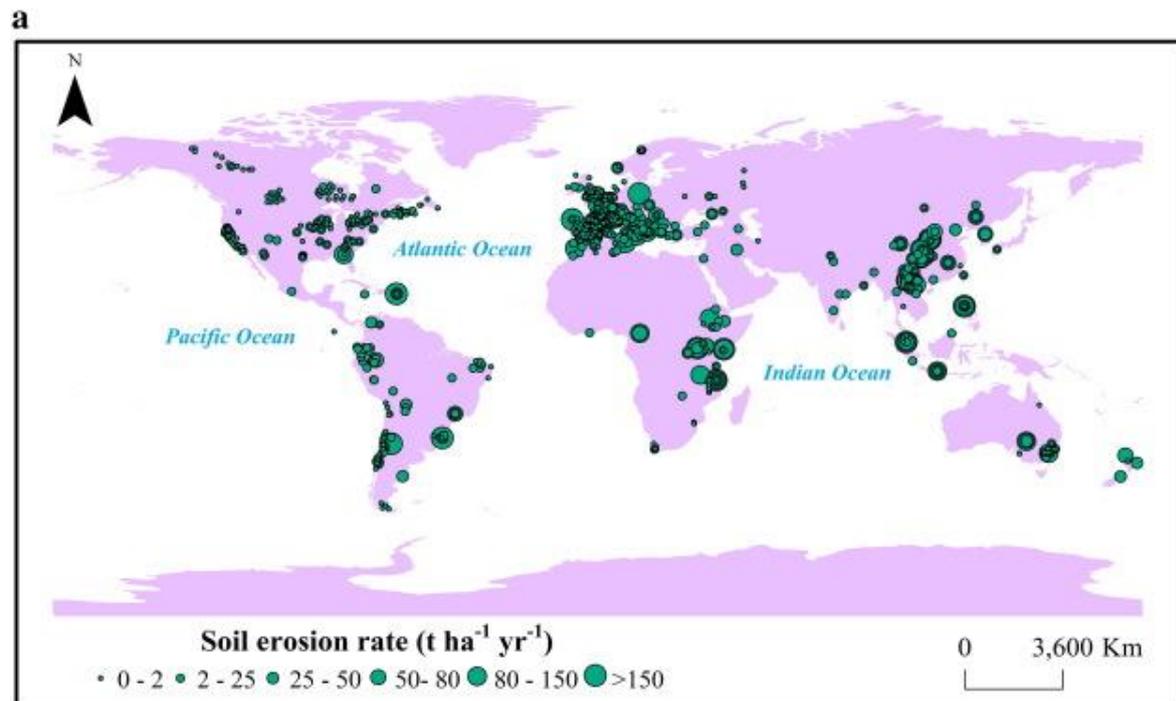


Figure 4.3. Map of observed soil erosion rates in database of 4377 entries by García-Ruiz et al., 2015). The map was published by (Li and Fang 2016).

An important finding from that database is that almost any erosion rate is possible under almost any climatic condition (García-Ruiz et al. 2015). Even if the results show few clear relationships between erosion and land conditions, the authors highlighted four observations: 1) the highest erosion rates were found in relation to agricultural activities – even though moderate erosion rates were also found in agricultural settings, 2) high erosion rates after forest fires were not observed (although the cases were few), 3) land covered by shrubs showed generally low erosion rates, 4) pasture land showed generally medium rates of erosion. Some important findings for the link between soil erosion and climate change can be noted from erosion measurements: erosion rates tend to increase with increasing mean annual rainfall, with a peak in the interval of 1000 to 1400 mm annual rainfall (García-Ruiz et al. 2015) (*low confidence*). However, such relationships are overshadowed by the fact that most rainfall events do not cause any erosion, instead erosion is caused by a few high intensity rainfall events (Fischer et al. 2016; Zhu et al. 2019). Hence mean annual rainfall is not a good predictor of erosion (Gonzalez-Hidalgo et al. 2012, 2009). In the context of climate change, it means the tendency of rainfall patterns to change towards more intensive precipitation events is serious. Such patterns have already been observed widely, even in cases where the total rainfall is decreasing (Trenberth 2011). The findings generally confirm the strong consensus about the importance of vegetation cover as a protection against soil erosion, emphasising how extremely important land management is for controlling erosion.

In the Mediterranean region, the observed and expected decrease in annual rainfall due to climate change is accompanied by an increase of rainfall intensity and hence erosivity (Capolongo et al. 2008). In tropical and sub-tropical regions, the on-site impacts of soil erosion dominate, and are manifested in very high rates of soil loss, in some cases exceeding $100 \text{ t ha}^{-1} \text{ yr}^{-1}$ (Tadesse 2001; García-Ruiz et al. 2015). In temperate regions, the off-site costs of soil erosion are often a greater concern, for example siltation of dams and ponds, downslope damage to property, roads and other infrastructure (Boardman 2010). In cases where water erosion occurs the down-stream effects, such as siltation of dams, are often significant and severe in terms of environmental and economic damages

1 (Kidane and Alemu 2015; Reinwarth et al. 2019; Quiñonero-Rubio et al. 2016; Adeogun et al. 2018;
2 Ben Slimane et al. 2016).

3 The distribution of wet and dry spells also affects land degradation although uncertainties remain
4 depending on resolution of climate models used for prediction (Kendon et al. 2014). Changes in
5 timing of rainfall events may have significant impacts on processes of soil erosion through changes in
6 wetting and drying of soils (Lado et al. 2004).

7 Soil moisture content is affected by changes in evapotranspiration and evaporation which may
8 influence the partitioning of water into surface and subsurface runoff (Li and Fang 2016; Nearing et
9 al. 2004b). This partitioning of rainfall can have a decisive effect on erosion (Stocking et al. 2001).

10 Wind erosion is a serious problem in agricultural regions, not only in drylands (Wagner 2013). Near-
11 surface wind speeds over land areas have decreased in recent decades (McVicar and Roderick 2010),
12 partly as a result of changing surface roughness (Vautard et al. 2010). Theoretically (Bakun 1990;
13 Bakun et al. 2015) and empirically (Sydeman et al. 2014; England et al. 2014) mean winds along
14 coastal regions worldwide have increased with climate change (*medium evidence, high agreement*).
15 Other studies of wind and wind erosion have not detected any long-term trend suggesting that climate
16 change has altered wind patterns outside drylands in a way that can significantly affect the risk of
17 wind erosion (Pryor and Barthelmie 2010; Barring et al. 2003). Therefore, the findings regarding wind
18 erosion and climate change are inconclusive, partly due to inadequate measurements.

19 Global mean temperatures are rising worldwide, but particularly in the Arctic region (*high*
20 *confidence*) (IPCC 2018a). Heat stress from extreme temperatures and heatwaves (multiple days of
21 hot weather in a row) have increased markedly in some locations in the last three decades (*high*
22 *confidence*), and are *virtually certain* to continue during the 21st century (Olsson et al. 2014a). The
23 IPCC Special Report on Global Warming of 1.5°C concluded that human-induced global warming has
24 already caused more frequent heatwaves in most of land regions, and that climate models project
25 robust differences between present-day and global warming up to 1.5°C and between 1.5°C and 2°C
26 (Hoegh-Guldberg et al. 2018). Direct temperature effects on soils are of two kinds. Firstly, permafrost
27 thawing leads to soil degradation in boreal and high altitude regions (Yang et al. 2010; Jorgenson and
28 Osterkamp 2005). Secondly, warming alters the cycling of nitrogen (N) and carbon (C) in soils, partly
29 due to impacts on soil microbiota (Solly et al. 2017). There are many studies with particularly strong
30 experimental evidence, but a full understanding of cause and effect is contextual and elusive (Conant
31 et al. 2011a,b; Wu et al. 2011). This is discussed comprehensively in Chapter 2.

32 Climate change, including increasing atmospheric CO₂ levels, affects vegetation structure and
33 function and hence conditions for land degradation. Exactly how vegetation responds to changes
34 remains a research task. In a comparison of seven global vegetation models under four representative
35 concentration pathways Friend et al., (2014) found that all models predicted increasing vegetation
36 carbon storage, however with substantial variation between models. An important insight compared
37 with previous understanding is that structural dynamics of vegetation seems to play a more important
38 role for carbon storage than vegetation production (Friend et al. 2014). The magnitude of CO₂
39 fertilisation of vegetation growth, and hence conditions for land degradation is still uncertain (Holtum
40 and Winter 2010), particularly in tropical rainforests (Yang et al. 2016). For more discussion on this
41 topic, see Chapter 2 in this report.

42 In summary, rainfall changes attributed to human-induced climate change have already intensified
43 drivers of land degradation (*robust evidence, high agreement*) but attributing land degradation to
44 climate change is challenging because of the importance of land management (*medium evidence, high*
45 *agreement*). Changes in climate variability modes, such as in monsoons and ENSO events, can also
46 affect land degradation (*low evidence, low agreement*).

1 **4.2.3.2 Indirect and complex linkages with climate change**

2 Many important indirect linkages between land degradation and climate change occur via agriculture,
3 particularly through changing outbreaks of pests (Rosenzweig et al. 2001; Porter et al. 1991; Thomson
4 et al. 2010; Dhanush et al. 2015; Lamichhane et al. 2015), which is covered comprehensively in
5 Chapter 5. More negative impacts have been observed than positive ones (IPCC 2014b). After 2050
6 the risk of yield losses increase as a result of climate change in combination with other drivers
7 (*medium confidence*) and such risks will increase dramatically if global mean temperatures increase
8 by ~4°C (*high confidence*) (Porter et al. 2014). The reduction (or plateauing) in yields in major
9 production areas (Brisson et al. 2010; Lin and Huybers 2012; Grassini et al. 2013) may trigger
10 cropland expansion elsewhere, either into natural ecosystems, marginal arable lands or intensification
11 on already cultivated lands, with possible consequences for increasing land degradation.

12 Precipitation and temperature changes will trigger changes in land- and crop management, such as
13 changes in planting and harvest dates, type of crops, and type of cultivars, which may alter the
14 conditions for soil erosion (Li and Fang 2016).

15 Much research has tried to understand how plants are affected by a particular stressor, for example
16 drought, heat, or waterlogging, including effects on belowground processes. But less research has
17 tried to understand how plants are affected by several simultaneous stressors – which of course is
18 more realistic in the context of climate change (Mittler 2006; Kerns et al. 2016) and from a hazards
19 point of view (see 7.2.1). From an attribution point of view, such a complex web of causality is
20 problematic if attribution is only done through statistical significant correlation. It requires a
21 combination of statistical links and theoretically informed causation, preferably integrated into a
22 model. Some modelling studies have combined several stressors with geomorphologically explicit
23 mechanisms (using the WEPP model) and realistic land use scenarios, and found severe risks of
24 increasing erosion from climate change (Mullan et al. 2012; Mullan 2013). Other studies have
25 included various management options, such as changing planting and harvest dates (Zhang and
26 Nearing 2005; Parajuli et al. 2016; Routschek et al. 2014; Nunes and Nearing 2011), type of cultivars
27 (Garbrecht and Zhang 2015), and price of crops (Garbrecht et al. 2007; O’Neal et al. 2005) to
28 investigate the complexity of how new climate regimes may alter soil erosion rates.

29 In summary, climate change increases the risk of land degradation both in terms of likelihood and
30 consequence but the exact attribution to climate change is challenging due to several confounding
31 factors. But since climate change exacerbates most degradation processes it is clear that unless land
32 management is improved, climate change will result in increasing land degradation (*very high*
33 *confidence*).

34 **4.2.4 Approaches to assessing land degradation**

35 In a review of different approaches and attempts to map global land degradation, Gibbs and Salmon
36 (2015) identified four main approaches to map the global extent of degraded lands: expert opinions
37 (Oldeman and van Lynden 1998; Dregne 1998; Reed 2005; Bot et al. 2000), satellite observation of
38 vegetation greenness (e.g., remote sensing of NDVI (Normalized Difference Vegetation Index), EVI
39 (Enhanced Vegetation Index), PPI (Plant Phenology Index) (Yengoh et al. 2015; Bai et al. 2008c; Shi
40 et al. 2017a; Abdi et al. 2019; JRC 2018), biophysical models (biogeographical/ topological) (Cai et
41 al. 2011b; Hickler et al. 2005; Steinkamp and Hickler 2015; Stoorvogel et al. 2017) and inventories of
42 land use/condition. Together they provide a relatively complete evaluation, but none on its own
43 assesses the complexity of the process (Vogt et al. 2011; Gibbs and Salmon 2015). There is, however,
44 a robust consensus that remote sensing and field-based methods are critical to assess and monitor land
45 degradation, particularly over large areas (such as global, continental and sub-continental) although
46 there are still knowledge gaps to be filled (Wessels et al. 2007, 2004; Prince 2016; Ghazoul and
47 Chazdon 2017) as well as the problem of baseline (see 4.1.3).

1 Remote sensing can provide meaningful proxies of land degradation in terms of severity, temporal
2 development, and areal extent. These proxies of land degradation include several indexes that have
3 been used to assess land conditions and monitoring the changes of land condition, for example extent
4 of gullies, severe forms of rill and sheet erosion, and deflation. The presence of open-access, quality
5 controlled and continuously updated global databases of remote sensing data is invaluable, and is the
6 only method for consistent monitoring of large areas over several decades (Sedano et al. 2016; Brandt
7 et al. 2018b; Turner 2014). The NDVI, as a proxy for Net Primary Production (NPP, see glossary), is
8 one of the most commonly used methods to assess land degradation, since it indicates land cover, an
9 important factor for soil protection. Although NDVI is not a direct measure of vegetation biomass,
10 there is a close coupling between NDVI integrated over a season and in situ NPP (*high agreement,*
11 *robust evidence*) (see Higginbottom et al. 2014; Andela et al. 2013; Wessels et al. 2012).

12 Distinction between land degradation/improvement and the effects of climate variation is an important
13 and contentious issue (Murthy and Bagchi 2018; Ferner et al. 2018). There is no simple and
14 straightforward way to disentangle these two effects. The interaction of different determinants of
15 primary production is not well understood. A key barrier to this is a lack of understanding of the
16 inherent inter-annual variability of vegetation (Huxman et al. 2004; Knapp and Smith 2001; Ruppert
17 et al. 2012; Bai et al. 2008a; Jobbágy and Sala 2000). One possibility is to compare potential land
18 productivity modelled by vegetation models and actual productivity measured by remote sensing
19 (Seaquist et al. 2009; Hickler et al. 2005; van der Esch et al. 2017), but the difference in spatial
20 resolution, typically 0.5 degrees for vegetation models compared to 0.25–0.5 km for remote sensing
21 data, is hampering the approach. Moderate-resolution Imaging Spectroradiometer, or MODIS,
22 provides higher spatial resolution (up to 0.25 km), delivers data for the Enhanced Vegetation Index
23 (EVI) which is calculated similarly to NDVI and have showed robust approach to estimate spatial
24 patterns of global annual primary productivity (Shi et al. 2017b; Testa et al. 2018).

25 Another approach to disentangle the effects of climate and land use/management is to use the Rain
26 Use Efficiency (RUE), defined as the biomass production per unit of rainfall, as an indicator (Le
27 Houerou 1984; Prince et al. 1998; Fensholt et al. 2015). A variant of the RUE approach is the residual
28 trend (RESTREND) of a NDVI time-series, defined as the fraction of the difference between the
29 observed NDVI and the NDVI predicted from climate data (Yengoh et al. 2015; John et al. 2016).
30 These two metrics aim to estimate the NPP, rainfall and the time dimensions. They are simple
31 transforms of the same three variables: RUE shows the NPP relationship with rainfall for individual
32 years, while RESTREND is the interannual change of RUE; also, both consider that rainfall is the
33 only variable that affects biomass production. They are legitimate metrics when used appropriately,
34 but in many cases they involve oversimplifications and yield misleading results (Fensholt et al. 2015;
35 Prince et al. 1998).

36 Furthermore, increases in NPP do not always indicate improvement in land condition/reversal of land
37 degradation, since this does not account for changes in vegetation composition. It could, for example,
38 result from conversion of native forest to plantation, or due to bush encroachment, which many
39 consider to be a form of land degradation (Ward 2005). Also, NPP may be increased by irrigation,
40 which can enhance productivity in the short-medium term while increasing risk of soil salinisation in
41 the long term (Niedertscheider et al. 2016).

42 Recent progress and expanding time series of canopy characterisations based on passive microwave
43 satellite sensors have offered rapid progress in regional and global descriptions of forest degradation
44 and recovery trends (Tian et al. 2017). The most common proxy is VOD (vertical optical depth) and
45 has already been used to describe global forest/savannah carbon stock shifts over two decades,
46 highlighting strong continental contrasts (Liu et al. 2015a) demonstrating the value of this approach to
47 monitor forest degradation at large scales. Contrasting NDVI which is only sensitive to vegetation
48 “greenness”, from which primary production can be modelled, VOD is also sensitive to water in

1 woody parts of the vegetation and hence provides a view of vegetation dynamics that can be
2 complementary to NDVI. As well as the NDVI, VOD also needs to be corrected to take into account
3 the rainfall variation (Andela et al. 2013).

4 Even though remote sensing offers much potential, its application to land degradation and recovery
5 remains challenging as structural changes often occur at scales below the detection capabilities of
6 most remote sensing technologies. Additionally, if the remote sensing is based on vegetation indexes
7 data, other forms of land degradation, such as nutrient depletion, changes of soil physical or biological
8 properties, loss of values for humans, among others, cannot be inferred indirectly by remote sensing.
9 The combination of remotely sensed images and field based approach can give improved estimates of
10 carbon stocks and tree biodiversity (Imai et al. 2012; Fujiki et al. 2016).

11 Additionally, the majority of trend techniques employed would be capable of detecting only the most
12 severe of degradation processes, and would therefore not be useful as a degradation early-warning
13 system (Higginbottom et al. 2014; Wessels et al. 2012). However, additional analyses using higher
14 resolution imagery, such as the Landsat and SPOT satellites, would be well suited to provide further
15 localized information on trends observed (Higginbottom et al. 2014). New approaches to assess land
16 degradation using high spatial resolution are developing but the need for time series makes progress
17 slow. The use of synthetic aperture radar (SAR) data has been shown to be advantageous for the
18 estimation of soil surface characteristics, in particular surface roughness and soil moisture (Gao et al.
19 2017; Bousbih et al. 2017), and detecting and quantifying selective logging (Lei et al. 2018). It is still
20 necessary to maintain the efforts to fully assess land degradation using remote sensing.

21 Computer simulation models can be used alone or combined with the remote sensing observations to
22 assess land degradation. The RUSLE (Revised Universal Soil Loss Equation) can be used, to some
23 extent, to predict the long-term average annual soil loss by water erosion. RUSLE has been constantly
24 revisited to estimate soil loss based on the product of rainfall–runoff erosivity, soil erodibility, slope
25 length and steepness factor, conservation factor, and support practice parameter (Nampak et al. 2018).
26 Inherent limitations of RUSLE include data-sparse regions, inability to account for soil loss from
27 gully erosion or mass wasting events, and that it does not predict sediment pathways from hillslopes
28 to water bodies (Benavidez et al. 2018). Since RUSLE models only provide gross erosion, the
29 integration of a further module in the RUSLE scheme to estimate the sediment yield from the
30 modelled hillslopes is needed. The spatially distributed sediment delivery model WaTEM/SEDEM
31 has been widely tested in Europe (Borrelli et al. 2018). Wind erosion is another factor that needs to be
32 taken into account in the modelling of soil erosion (Webb et al. 2017a, 2016). Additional models need
33 to be developed to include the limitations of the RUSLE models.

34 Regarding the field based approach to assess land degradation, there are multiple indicators that
35 reflect functional ecosystem processes linked to ecosystem services and, thus, to the value for
36 humans. These indicators are a composite set of measurable attributes from different factors, such as
37 climate, soil, vegetation, biomass, management, among others, that can be used together or to develop
38 indexes to better assess land degradation (Allen et al. 2011; Kosmas et al. 2014).

39 Declines in vegetation cover, changes in vegetation structure, decline in mean species abundances,
40 decline in habitat diversity, changes in abundance of specific indicator species, reduced vegetation
41 health and productivity, and vegetation management intensity and use, are the most common
42 indicators in the vegetation condition of forest and woodlands (Stocking et al. 2001; Wiesmair et al.
43 2017; Ghazoul and Chazdon 2017; Alkemade et al. 2009).

44 Several indicators of the soil quality (soil organic matter, depth, structure, compaction, texture, pH,
45 C:N ratio, aggregate size distribution and stability, microbial respiration, soil organic carbon,
46 salinisation, among others) have been proposed (see also 2.2) (Schoenholtz et al. 2000). Among these,
47 soil organic matter (SOM) directly and indirectly drives the majority of soil functions. Decreases in

1 SOM can lead to a decrease in fertility and biodiversity, as well as a loss of soil structure, causing
2 reductions in water holding capacity, increased risk of erosion (both wind and water) and increased
3 bulk density and hence soil compaction (Allen et al. 2011; Certini 2005; Conant et al. 2011a). Thus,
4 indicators related with the quantity and quality of the SOM are necessary to identify land degradation
5 (Pulido et al. 2017; Dumanski and Pieri 2000). The composition of the microbial community is *very*
6 *likely* to be positive impacted by both climate change and land degradation processes (Evans and
7 Wallenstein 2014; Wu et al. 2015; Classen et al. 2015), thus changes in microbial community
8 composition can be very useful to rapidly reflect land degradation (e.g. forest degradation increased
9 the bacterial alpha-diversity indexes) (Flores-Rentería et al. 2016; Zhou et al. 2018). These indicators
10 might be used as a set of indicators site-dependent, and in a plant-soil system (Ehrenfeld et al. 2005).

11 Useful indicators of degradation and improvement include changes in ecological processes and
12 disturbance regimes that regulate the flow of energy and materials and that control ecosystem
13 dynamics under a climate change scenario. Proxies of dynamics include spatial and temporal turnover
14 of species and habitats within ecosystems (Ghazoul et al. 2015; Bahamondez and Thompson 2016).
15 Indicators in agricultural lands include crop yield decreases and difficulty in maintaining yields
16 (Stocking et al. 2001). Indicators of landscape degradation/improvement in fragmented forest
17 landscapes include the extent, size, and distribution of remaining forest fragments, an increase in edge
18 habitat, and loss of connectivity and ecological memory (Zahawi et al. 2015; Pardini et al. 2010).

19 In summary, as land degradation is such a complex and global process there is no single method by
20 which land degradation can be estimated objectively and consistently over large areas (*very high*
21 *confidence*). However, many approaches exist that can be used to assess different aspects of land
22 degradation or provide proxies of land degradation. Remote sensing, complemented by other kinds of
23 data (i.e., field observations, inventories, expert opinions), is the only method that can generate
24 geographically explicit and globally consistent data over time scales relevant for land degradation
25 (several decades).

26 **4.3 Status and current trends of land degradation**

27 The scientific literature on land degradation often excludes forest degradation, yet here we attempt to
28 assess both issues. Because of the different bodies of scientific literature, we assess land degradation
29 and forest degradation under different sub-headings, and where possible draw integrated conclusions.

30 **4.3.1 Land degradation**

31 There are no reliable global maps of the extent and severity of land degradation (Gibbs and Salmon
32 2015; Prince et al. 2018; van der Esch et al. 2017), despite the fact that land degradation is a severe
33 problem (Turner et al. 2016). The reasons are both conceptual, i.e., how is land degradation defined,
34 using what baseline (Herrick et al. 2019) or over what time period, and methodological, i.e. how can it
35 be measured (Prince et al. 2018). Although there is a strong consensus that land degradation is a
36 reduction in productivity of the land or soil, there are diverging views regarding the spatial and
37 temporal scales at which land degradation occurs (Warren 2002), and how this can be quantified and
38 mapped. Proceeding from the definition in this report, there are also diverging views concerning
39 ecological integrity and the value to humans. A comprehensive treatment of the conceptual discussion
40 about land degradation is provided by the recent report on land degradation from the
41 Intergovernmental Science-Policy Platform on Biodiversity and Ecosystem Services, IPBES
42 (Montanarella et al. 2018).

43 A review of different attempts to map global land degradation, based on expert opinion, satellite
44 observations, biophysical models and a data base of abandoned agricultural lands, suggested that

1 between <10 M km² to 60 M km² (corresponding to 8–45% of the ice-free land area) have been
2 degraded globally (Gibbs and Salmon, 2015) (*very low confidence*).

3 One often used global assessment of land degradation used trends in NDVI as a proxy for land
4 degradation and improvement during the period 1983 to 2006 (Bai et al. 2008b,c) with an update to
5 2011 (Bai et al. 2015). These studies, based on very coarse resolution satellite data (8 km NOAA
6 AVHRR), indicated that between 22% and 24% of the global ice-free land area was subject to a
7 downward trend, while about 16% showed an increasing trend. The study also suggested, contrary to
8 earlier assessments (Middleton and Thomas 1997), that drylands were not among the most affected
9 regions. Another study using a similar approach for the period 1981-2006 suggested that about 29%
10 of the global land area is subject to ‘land degradation hotspots’, i.e. areas with acute land degradation
11 in need of particular attention. These hotspot areas were distributed over all agro-ecological regions
12 and land cover types. Two different studies have tried to link land degradation, identified by NDVI as
13 a proxy, and number of people affected: Le et al. (2016) estimated that at least 3.2 billion people were
14 affected, while Barbier and Hochard (2016, 2018) estimated that 1.33 billion people were affected, of
15 which 95% were living in developing countries.

16 Yet another study, using a similar approach and type of remote sensing data, compared NDVI trends
17 with biomass trends calculated by a global vegetation model over the period 1982-2010 and found
18 that 17-36% of the land areas showed a negative NDVI trend while a positive or neutral trend was
19 predicted in modelled vegetation (Schut et al. 2015). The World Atlas of Desertification (3rd edition)
20 includes a global map of land productivity change over the period 1999 to 2013, which is one useful
21 proxy for land degradation (Cherlet et al. 2018). Over that period about 20% of the global ice-free
22 land area shows signs of declining or unstable productivity, whereas about 20% shows increasing
23 productivity. The same report also summarized the productivity trends by land categories and found
24 that most forest land showed increasing trends in productivity while rangelands had more declining
25 trends than increasing trends (Fig 4.4). These productivity assessments, however, do not distinguish
26 between trends due to climate change and trends due to other factors. A recent analysis of “greening”
27 of the world using MODIS time series of NDVI 2000 – 2017, shows a striking increase in the
28 greening over China and India. In China the greening is seen over both forested areas, 42%, and
29 cropland areas, in which 32% is increasing (see Section 4.9.3). In India, the greening is almost
30 entirely associated with cropland (82%) (Chen et al. 2019).

31 All the studies of vegetation trends referred to above show that there are regionally-differentiated
32 trends of either decreasing or increasing vegetation. When comparing vegetation trends with trends in
33 climatic variables, Schut et al. (2015) found very few areas (1-2%) where an increase in vegetation
34 trend was independent of the climate drivers, and that study suggested that positive vegetation trends
35 are primarily caused by climatic factors.

36



1

2

3

4

5

6

Figure 4.4. Proportional global land productivity trends by land cover/land use class. (Cropland includes arable land, permanent crops and mixed classes with over 50% crops; Grassland includes natural grassland and managed pasture land; Rangelands include shrub land, herbaceous and sparsely vegetated areas; Forest land includes all forest categories and mixed classes with tree cover greater than 40%). Data source: Copernicus Global Land SPOT VGT, 1999-2013.

7

8

9

10

11

12

13

In an attempt to go beyond the mapping of global vegetation trends for assessing land degradation, Borelli et al. (2017) used a soil erosion model (RUSLE) and suggested that soil erosion is mainly caused in areas of crop land expansion, particularly in sub-Saharan Africa, South America and Southeast Asia. The method is controversial for both conceptual reasons (i.e., the ability of the model to capture the most important erosion processes) and data limitations (i.e., the availability of relevant data at regional to global scales), and its validity for assessing erosion over large areas has been questioned by several studies (Baveye 2017; Evans and Boardman 2016a,b; Labrière et al. 2015).

14

15

16

17

18

19

20

21

22

23

24

25

26

27

An alternative to using remote sensing for assessing the state of land degradation is to compile field based data from around the globe (Turner et al. 2016). In addition to the problems of definitions and baselines, this approach is also hampered by the lack of standardized methods used in the field. An assessment of the global severity of soil erosion in agriculture, based on 1,673 measurements around the world (compiled from 201 peer reviewed articles), indicated that the global net median rate of soil formation (i.e., formation minus erosion) is about 0.004 mm yr^{-1} ($\sim 0.05 \text{ t ha}^{-1}\text{yr}^{-1}$) compared with the median net rate of soil loss in agricultural fields, 1.52 mm yr^{-1} ($\sim 18 \text{ t ha}^{-1}\text{yr}^{-1}$) in tilled fields and 0.065 mm yr^{-1} ($\sim 0.8 \text{ t ha}^{-1}\text{yr}^{-1}$) in no-till fields (Montgomery 2007a). This means that the rate of soil erosion from agricultural fields is in between 380 (conventional tilling) and 16 times (no-till) the natural rate of soil formation (*medium agreement, limited evidence*). These approximate figures are supported by another large meta-study including over 4000 sites around the world (see Figure 4.4) where the average soil loss from agricultural plots was $\sim 21 \text{ t ha}^{-1}\text{yr}^{-1}$ (García-Ruiz et al. 2015). Climate change, mainly through the intensification of rainfall, will further increase these rates unless land management is improved (*high agreement, medium evidence*).

1 Soils contain about 1500 Gt of organic carbon (median across 28 different estimates presented by
2 (Scharlemann et al. 2014)), which is about 1.8 times more carbon than in the atmosphere (Ciais et al.
3 2013) and 2.3 – 3.3 times more than what is held in the terrestrial vegetation of the world (Ciais et al.
4 2013). Hence, land degradation including land conversion leading to soil carbon losses has the
5 potential to impact the atmospheric concentration of CO₂ substantially. When natural ecosystems are
6 cultivated they lose soil carbon that accumulated over long time periods. The loss rate depends on the
7 type of natural vegetation and how the soil is managed. Estimates of the magnitude of loss vary but
8 figures between 20% and 59% have been reported in several meta studies (Poeplau and Don 2015;
9 Wei et al. 2015; Li et al. 2012; Murty et al. 2002; Guo and Gifford 2002). The amount of soil carbon
10 lost explicitly due to land degradation after conversion is hard to assess due to large variation in local
11 conditions and management, see also Chapter 2.

12 From a climate change perspective, land degradation plays an important role in the dynamics of
13 nitrous oxide (N₂O) and methane (CH₄). N₂O is produced by microbial activity in the soil and the
14 dynamics are related to both management practices and weather conditions while CH₄ dynamics are
15 primarily determined by the amount of soil carbon and to what extent the soil is subject to water
16 logging (Palm et al. 2014), see also Chapter 2.

17 Several attempts have been made to map the human footprint on the planet (Čuček et al. 2012; Venter
18 et al. 2016) but they in some cases confuse human impact on the planet with degradation. From our
19 definition it is clear that human impact (or pressure) is not synonymous with degradation but
20 information on the human footprint provides a useful mapping of potential non-climatic drivers of
21 degradation.

22 In summary, there are no uncontested maps of the location, extent and severity of land degradation.
23 Proxy estimates based on remote sensing of vegetation dynamics provide one important information
24 source, but attribution of the observed changes in productivity to climate change, human activities, or
25 other drivers is hard. Nevertheless, the different attempts to map the extent of global land degradation
26 using remotely sensed proxies show some convergence and suggest that about a quarter of the ice free
27 land area is subject to some form of land degradation (*limited evidence, medium agreement*) affecting
28 about 3.2 billion people (*low confidence*). Attempts to estimate the severity of land degradation
29 through soil erosion estimates suggest that soil erosion is a serious form of land degradation in
30 croplands closely associated with unsustainable land management in combination with climatic
31 parameters, some of which are subject to climate change (*limited evidence, high agreement*). Climate
32 change is one among several causal factors in the status and current trends of land degradation
33 (*limited evidence, high agreement*).

34 **4.3.2 Forest degradation**

35 Quantifying degradation in forests has also proven difficult. Indicators that remote sensing or
36 inventory methods can measure more easily than reductions in biological productivity, losses of
37 ecological integrity or value to humans include reductions in canopy cover or carbon stocks.
38 However, the causes of reductions in canopy cover or carbon stocks can be many (Curtis et al. 2018),
39 including natural disturbances (e.g., fires, insects and other forest pests), direct human activities (e.g.,
40 harvest, forest management) and indirect human impacts (such as climate change) and these may not
41 reduce long-term biological productivity. In many boreal, some temperate and other forest types
42 natural disturbances are common, and consequently these disturbance-adapted forest types are
43 comprised of a mosaic of stands of different ages and stages of stand recovery following natural
44 disturbances. In those managed forests where natural disturbances are uncommon or suppressed,
45 harvesting is the primary determinant of forest age-class distributions.

46 Quantifying forest degradation as a reduction in productivity, carbon stocks or canopy cover also
47 requires that an initial condition (or baseline) is established against which this reduction is assessed

1 (see Section 4.1.4). In forest types with rare stand-replacing disturbances, the concept of “intact” or
2 “primary” forest has been used to define the initial condition (Potapov et al. 2008) but applying a
3 single metric can be problematic (Bernier et al. 2017). Moreover, forest types with frequent stand-
4 replacing disturbances, such as wildfires, or with natural disturbances that reduce carbon stocks, such
5 as some insect outbreaks, experience over time a natural variability of carbon stocks or canopy
6 density making it more difficult to define the appropriate baseline carbon density or canopy cover
7 against which to assess degradation. In these systems, forest degradation cannot be assessed at the
8 stand level, but requires a landscape-level assessment that takes into consideration the stand age-class
9 distribution of the landscape, which reflects natural and human disturbance regimes over past decades
10 to centuries and also considers post-disturbance regrowth (van Wagner 1978; Volkova et al. 2018;
11 Lorimer and White 2003).

12 The lack of a consistent definition of forest degradation also affects the ability to establish estimates
13 of the rates or impacts of forest degradation because the drivers of degradation are not clearly defined
14 (Sasaki and Putz 2009). Moreover, the literature at times confounds estimates of forest degradation
15 and deforestation (i.e. the conversion of forest to non-forest land uses). Deforestation is a change in
16 land use, while forest degradation is not, although severe forest degradation can ultimately lead to
17 deforestation.

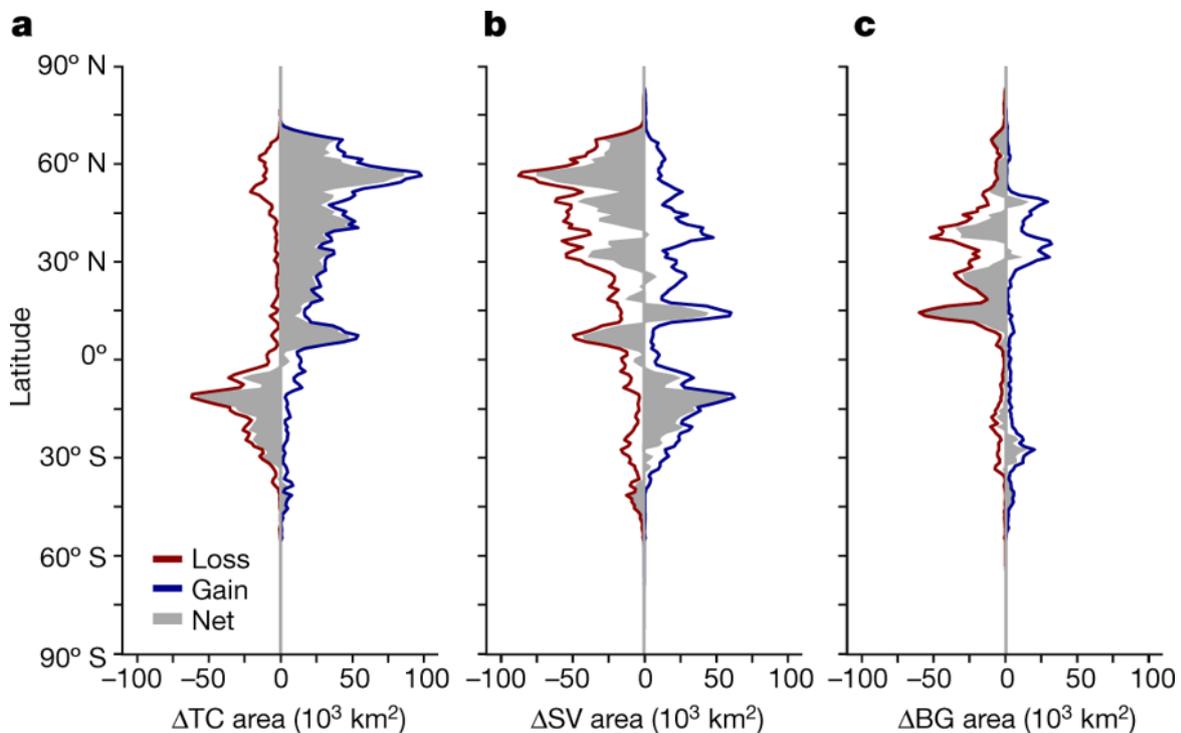
18 Based on empirical data provided by 46 countries, the drivers for deforestation (due to commercial
19 agriculture) and forest degradation (due to timber extraction and logging) are similar in Africa, Asia
20 and Latin America (Hosonuma et al. 2012). More recently, global forest disturbance over the period
21 2001 – 2015 was attributed to commodity driven deforestation ($27 \pm 5\%$), forestry ($26 \pm 4\%$), shifting
22 agriculture ($24 \pm 3\%$) and wildfire ($23 \pm 4\%$). The remaining $0.6 \pm 0.3\%$ was attributed to the
23 expansion of urban centers (Curtis et al. 2018).

24 The trends of productivity shown by several remote sensing studies (see previous section) are largely
25 consistent with mapping of forest cover and change using a 34 year time series of coarse resolution
26 satellite data (NOAA AVHRR) (Song et al. 2018). This study, based on a thematic classification of
27 satellite data, suggests that (i) global tree canopy cover increased by 2.24 million km² between 1982
28 and 2016 (corresponding to +7.1%) but with regional differences that contribute a net loss in the
29 tropics and a net gain at higher latitudes, and (ii) the fraction of bare ground decreased by 1.16 million
30 km² (corresponding to -3.1%), mainly in agricultural regions of Asia (Song et al. 2018), see Figure
31 4.5. Other tree or land cover datasets show opposite global net trends (Li et al. 2018b), but high
32 agreement in terms of net losses in the tropics and large net gains in the temperate and boreal zones
33 (Li et al. 2018b; Song et al. 2018; Hansen et al. 2013). Differences across global estimates are further
34 discussed in Chapter 1 (1.1.2.3) and Chapter 2.

35 The changes detected from 1982 to 2016 were primarily linked to direct human action, such as land-
36 use changes (about 60% of the observed changes), but also to indirect effects, such as human induced
37 climate change (about 40% of the observed changes) (Song et al. 2018), a finding also supported by a
38 more recent study (Chen et al. 2019). The climate induced effects were clearly discernible in some
39 regions, such as forest decline in the US Northwest due to increasing pest infestation and increasing
40 fire frequency (Lesk et al. 2017; Abatzoglou and Williams 2016; Seidl et al. 2017), warming induced
41 vegetation increase in the Arctic region, general greening in the Sahel probably as a result of
42 increasing rainfall and atmospheric CO₂, and advancing treelines in mountain regions (Song et al.
43 2018).

44 Keenan et al.(Keenan et al. 2015) and Sloan and Sayer (2015) studied the 2015 Forest Resources
45 Assessment (FRA) of the FAO (FAO 2016) and found that the total forest area from 1990 to 2015
46 declined by 3%, an estimate that is supported by a global remote sensing assessment of forest area
47 change that found a 2.8% decline between 1990-2010 (D’Annunzio et al. 2017; Lindquist and
48 D’Annunzio 2016). The trend in deforestation is, however, contradicting between these two global

1 assessments with FAO (2016) suggesting deforestation is slowing down while the remote sensing
 2 assessments finds it to be accelerating (D'Annunzio et al. 2017). Recent estimates (Song et al. 2018)
 3 owing to semantic and methodological differences (see Chapter 1, section 1.1.2.3) suggest global tree
 4 cover to have increased over the period 1982-2016, which contradicts the forest area dynamics
 5 assessed by FAO (2016, Lindquist and D'Annunzio 2016). The loss rate in tropical forest areas from
 6 2010 to 2015 is 55 000 km² yr⁻¹. According to the FRA the global natural forest area also declined
 7 from 39.61 M km² to 37.21 M km² during the period 1990 to 2015 (Keenan et al. 2015).



8

9 **Figure 4.5. Diagrams showing latitudinal profiles of land cover change over the period 1982 to 2016 based**
 10 **on analysis of time-series of NOAA AVHRR imagery: a, Tree canopy cover change (ΔTC). b, Short**
 11 **vegetation cover change (ΔSV). c, Bare ground cover change (ΔBG). Area statistics were calculated for**
 12 **every 1° of latitude (Song et al. 2018). Source of data: NOAA AVHRR.**

13 Since 1850, deforestation globally contributed 77% of the emissions from land-use and land-cover
 14 change (LULCC) while degradation contributed 10% (with the remainder originating from non-forest
 15 land uses) (Houghton and Nassikas 2018). That study also showed large temporal and regional
 16 differences with northern mid-latitude forests currently contributing carbon sinks due to increasing
 17 forest area and forest management. However, the contribution to carbon emissions of degradation as
 18 percentage of total forest emissions (degradation and deforestation) are uncertain, with estimates
 19 varying from 25% (Pearson et al. 2017) to nearly 70% of carbon losses (Baccini et al. 2017). The 25%
 20 estimate refers to an analysis of 74 developing countries within tropical and subtropical regions
 21 covering 22 million km² for the period 2005-2010 while the 70% estimate refers to an analysis of the
 22 tropics for the period 2003-2014, but by and large the scope of these studies is the same. Pearson et al.
 23 (2017) estimated annual gross emissions of 2.1 Gt CO₂, of which 53% were derived from timber
 24 harvest, 30% from wood fuel harvest and 17% from forest fire. Estimating gross emissions only,
 25 creates a distorted representation of human impacts on the land sector carbon cycle. While forest
 26 harvest for timber and fuel wood and land-use change (deforestation) contribute to gross emissions, to
 27 quantify impacts on the atmosphere it is necessary to estimate net emissions, i.e. the balance of gross
 28 emissions and gross removals of carbon from the atmosphere through forest regrowth (Chazdon et al.
 29 2016a; Poorter et al. 2016; Sanquetta et al. 2018).

1 Current efforts to reduce atmospheric CO₂ concentrations can be supported by reductions in forest-
2 related carbon emissions and increases in sinks, which requires that the net impact of forest
3 management on the atmosphere be evaluated (Griscom et al. 2017). Forest management and the use
4 of wood products in GHG mitigation strategies result in changes in forest ecosystem C stocks,
5 changes in harvested wood product C stocks, and potential changes in emissions resulting from the
6 use of wood products and forest biomass that substitute for other emissions-intensive materials such
7 as concrete, steel and fossil fuels (Kurz et al. 2016; Lemprière et al. 2013; Nabuurs et al. 2007). The
8 net impact of these changes on GHG emissions and removals, relative to a scenario without forest
9 mitigation actions needs to be quantified, (e.g. Werner et al. 2010; Smyth et al. 2014; Xu et al. 2018).
10 Therefore, reductions in forest ecosystem C stocks alone are an incomplete estimator of the impacts of
11 forest management on the atmosphere (Nabuurs et al. 2007; Lemprière et al. 2013; Kurz et al. 2016;
12 Chen et al. 2018b). The impacts of forest management and the carbon storage in long-lived products
13 and landfills vary greatly by region, however, because of the typically much shorter life-span of wood
14 products produced from tropical regions compared to temperate and boreal regions (Earles et al. 2012;
15 Lewis et al. 2019; Jordan et al. 2018) (see also section 4.8.4).

16 Assessments of forest degradation based on remote sensing of changes in canopy density or land
17 cover, (e.g., (Hansen et al. 2013; Pearson et al. 2017) quantify changes in aboveground biomass C
18 stocks and require additional assumptions or model-based analyses to also quantify the impacts on
19 other ecosystem carbon pools including belowground biomass, litter, woody debris and soil carbon.
20 Depending on the type of disturbance, changes in aboveground biomass may lead to decreases or
21 increases in other carbon pools, for example, windthrow and insect induced tree mortality may result
22 in losses in aboveground biomass that are (initially) off-set by corresponding increases in dead
23 organic matter carbon pools (Yamanoi et al. 2015; Kurz et al. 2008), while deforestation will reduce
24 the total ecosystem carbon pool (Houghton et al. 2012).

25 A global study of current vegetation carbon stocks (450 Gt C), relative to a hypothetical condition
26 without land-use (916 Gt C), attributed 42-47% of carbon stock reductions to land management
27 effects without land-use change, while the remaining 53-58% of carbon stock reductions were
28 attributed to deforestation and other land-use changes (Erb et al. 2018). While carbon stocks in
29 European forests are lower than hypothetical values in the complete absence of human land use, forest
30 area and carbon stocks have been increasing over recent decades (McGrath et al. 2015; Kauppi et al.
31 2018). Studies of Gingrich et al. (2015) on the long-term trends in land-use over nine European
32 countries (Albania, Austria, Denmark, Germany, Italy, the Netherlands, Romania, Sweden and the
33 United Kingdom) also show an increase in forest land and reduction in cropland and grazing land
34 from the 19th century to the early 20th century. However, the extent to which human activities have
35 affected the productive capacity of forest lands is poorly understood. Biomass Production Efficiency
36 (BPE), i.e. the fraction of photosynthetic production used for biomass production, was significantly
37 higher in managed forests (0.53) compared to natural forests (0.41) (and it was also higher in
38 managed (0.44) compared to natural (0.63) grasslands) (Capioli et al. 2015). Managing lands for
39 production may involve trade-offs. For example, a larger proportion of Net Primary Production in
40 managed forests is allocated to biomass carbon storage, but lower allocation to fine roots is
41 hypothesised to reduce soil C stocks in the long-term (Noormets et al. 2015). Annual volume
42 increment in Finnish forests has more than doubled over the last century, due to increased growing
43 stock, improved forest management and environmental changes (Henttonen et al. 2017).

44 As economies evolve, the patterns of land-use and C stock changes associated with human expansion
45 into forested areas often include a period of rapid decline of forest area and carbon stocks, recognition
46 of the need for forest conservation and rehabilitation, and a transition to more sustainable land
47 management that is often associated with increasing carbon stocks, (e.g. Birdsey et al. 2006).
48 Developed and developing countries around the world are in various stages of forest transition

1 (Kauppi et al. 2018; Meyfroidt and Lambin 2011). Thus, opportunities exist for sustainable forest
2 management to contribute to atmospheric carbon targets through reduction of deforestation and
3 degradation, forest conservation, forest restoration, intensification of management, and enhancements
4 of carbon stocks in forests and harvested wood products (Griscom et al. 2017) (*medium evidence,*
5 *medium agreement*).

6 **4.4 Projections of land degradation in a changing climate**

7 Land degradation will be affected by climate change in both direct and indirect ways, and land
8 degradation will to some extent also feed-back into the climate system. The direct impacts are those in
9 which climate and land interact directly in time and space. Examples of direct impacts are when
10 increasing rainfall intensity exacerbates soil erosion, or when prolonged droughts reduce the
11 vegetation cover of the soil making it more prone to erosion and nutrient depletion. The indirect
12 impacts are those where climate change impacts and land degradation are separated in time and/or
13 space. Examples of such impacts are when declining agricultural productivity due to climate change
14 drives an intensification of agriculture elsewhere, which may cause land degradation. Land
15 degradation, if sufficiently widespread, may also feed back into the climate system by reinforcing
16 ongoing climate change.

17 Although climate change is exacerbating many land degradation processes (*high to very high*
18 *confidence*), prediction of future land degradation is challenging because land management practices
19 determine to a very large extent the state of the land. Scenarios of climate change in combination with
20 land degradation models can provide useful knowledge on what kind and extent of land management
21 will be necessary to avoid, reduce and reverse land degradation.

22 **4.4.1 Direct impacts on land degradation**

23 There are two main levels of uncertainty in assessing the risks of future climate change induced land
24 degradation. The first level, where uncertainties are comparatively low, is the changes of the
25 degrading agent, such as erosive power of precipitation, heat stress from increasing temperature
26 extremes (HÜVE et al. 2011), water stress from droughts, and high surface wind speed. The second
27 level of uncertainties, and where the uncertainties are much larger, relates to the above and
28 belowground ecological changes as a result of changes in climate, such as rainfall, temperature, and
29 increasing level of CO₂. Vegetation cover is crucial to protect against erosion (Mullan et al. 2012;
30 García-Ruiz et al. 2015).

31 Changes in rainfall patterns, such as distribution in time and space, and intensification of rainfall
32 events will increase the risk of land degradation, both in terms of likelihood and consequences (*high*
33 *agreement, medium evidence*). Climate induced vegetation changes will increase the risk of land
34 degradation in some areas (where vegetation cover will decline) (*medium confidence*). Landslides are
35 a form of land degradation that is induced by extreme rainfall events. There is a strong theoretical
36 reason for increasing landslide activity due to intensification of rainfall, but the empirical evidence is
37 so far lacking that climate change has contributed to landslides (Crozier 2010; Huggel et al. 2012;
38 Gariano and Guzzetti 2016), human disturbance may be a more important future trigger than climate
39 change (Froude and Petley 2018).

40 Erosion of coastal areas as a result of sea level rise will increase worldwide (*very high confidence*). In
41 cyclone prone areas (such as the Caribbean, Southeast Asia, and the Bay of Bengal) the combination
42 of sea level rise and more intense cyclones (Walsh et al. 2016b), and in some areas also land
43 subsidence (Yang et al. 2019; Shirzaei and Bürgmann 2018; Wang et al. 2018; Fuangswasdi et al.
44 2019; Keogh and Törnqvist 2019), will pose a serious risk to people and livelihoods (*very high*
45 *confidence*), in some cases even exceeding limits to adaption, see further section 4.8.4.1, 4.9.6, 4.9.8.

1 **4.4.1.1 Changes in water erosion risk due to precipitation changes**

2 The hydrological cycle is intensifying with increasing warming of the atmosphere. The intensification
3 means that the number of heavy rainfall events is increasing while the total number of rainfall events
4 tends to decrease (Trenberth 2011; Li and Fang 2016; Kendon et al. 2014; Guerreiro et al. 2018; Burt
5 et al. 2016a; Westra et al. 2014; Pendergrass and Knutti 2018) (*robust evidence, high agreement*).
6 Modelling of changes in land degradation as a result of climate change alone is hard because of the
7 importance of local contextual factors. As shown above, actual erosion rate is extremely dependent on
8 local conditions, primarily vegetation cover and topography (García-Ruiz et al. 2015). Nevertheless,
9 modelling of soil erosion risks has advanced substantially in recent decades and such studies are
10 indicative of future changes in the risk of soil erosion while actual erosion rates will still primarily be
11 determined by land management. In a review article, Li & Fang (Li and Fang 2016) summarised 205
12 representative modelling studies around the world where erosion models had been used in
13 combination with down-scaled climate models to assess future (between 2030 to 2100) erosion rates.
14 The meta-study by Li & Fang considered, where possible, climate change in terms of temperature
15 increase and changing rainfall regimes and their impacts on vegetation and soils. Almost all of the
16 sites had current soil loss rates above 1 t ha^{-1} (assumed to be the upper limit for acceptable soil erosion
17 in Europe) and 136 out of 205 studies predicted increased soil erosion rates. The percentage increase
18 in erosion rates varied between 1.2% to as much as over 1600%, whereas 49 out of 205 studies
19 projected more than 50% increase. Projected soil erosion rates varied substantially between studies
20 because the important of local factors, hence climate change impacts on soil erosion should preferably
21 be assessed at the local to regional scale, rather than the global (Li and Fang 2016).

22 Mesoscale convective systems (MCS), typically thunder storms, have increased markedly in recent 3-
23 4 decades in the USA and Australia and they are projected to increase substantially (Prein et al. 2017).
24 Using a climate model with the ability to represent MCS, Prein and colleagues were able to predict
25 future increases in frequency, intensity, and size of such weather systems. Findings include the 30%
26 decrease in number of MCS of $<40\text{ mm h}^{-1}$, but a sharp increase of 380% in the number of extreme
27 precipitation events of $>90\text{ mm h}^{-1}$ over the North American continent. The combined effect of
28 increasing precipitation intensity and increasing size of the weather systems implies that the total
29 amount of precipitation from these weather systems is expected to increase by up to 80% (Prein et al.
30 2017), which will substantially increase the risk of land degradation in terms of landslides, extreme
31 erosion events, flashfloods etc.

32 The potential impacts of climate change on soil erosion can be assessed by modelling the projected
33 changes in particular variables of climate change known to cause erosion, such as erosivity of rainfall.
34 A study of the conterminous United States based on three climate models and three scenarios (A2,
35 A1B, and B1) found that rainfall erosivity will increase in all scenarios, even if there are large spatial
36 differences – strong increase in NE and NW, and either weak or inconsistent trends in the SW and
37 mid-West (Segura et al. 2014).

38 In a study of how climate change will impact future soil erosion processes in the Himalayas, Gupta
39 and Kumar (2017) estimated that soil erosion will increase by about 27% in the near term (2020s) and
40 22% in the medium term (2080s), with little difference between scenarios. A study from Northern
41 Thailand estimated that erosivity will increase by 5% in the near term (2020s) and 14% in the medium
42 term (2080s), which would result in a similar increase of soil erosion, all other factors being constant
43 (Plangoen and Babel 2014). Observed rainfall erosivity has increased significantly in the lower Niger
44 Basin (Nigeria) and are predicted to increase further based on statistical downscaling of four General
45 Circulation Models (GCM) scenarios, with an estimated increase of 14%, 19% and 24% for the
46 2030s, 2050s, and 2070s respectively (Amanambu et al. 2019).

47 Many studies from around the world where statistical downscaling of GCM results have been used in
48 combination with process based erosion models show a consistent trend of increasing soil erosion

1 Using a comparative approach Serpa et al. (2015) studied two Mediterranean catchments (one dry and
2 one humid) using a spatially explicit hydrological model (SWAT) in combination with land use and
3 climate scenarios for 2071-2100. Climate change projections showed, on the one hand, decreased
4 rainfall and streamflow for both catchments whereas sediment export decreased only for the humid
5 catchment; projected land use change, from traditional to more profitable, on the other hand resulted
6 in increase in streamflow. The combined effect of climate and land use change resulted in reduced
7 sediment export for the humid catchment (-29% for A1B; -22% for B1) and increased sediment export
8 for the dry catchment (+222% for A1B; +5% for B1). Similar methods have been used elsewhere, also
9 showing the dominant effect of land use/land cover for runoff and soil erosion (Neupane and Kumar
10 2015).

11 A study of future erosion rates in Northern Ireland, using a spatially explicit erosion model in
12 combination with downscaled climate projections (with and without sub-daily rainfall intensity
13 changes), showed that erosion rates without land management changes would decrease by 2020s,
14 2050s and 2100s irrespective of changes in intensity, mainly as a result of a general decline in rainfall
15 (Mullan et al. 2012). When land management scenarios were added to the modelling, the erosion rates
16 started to vary dramatically for all three time periods, ranging from a decrease of 100% for no-till land
17 use, to an increase of 3621% for row crops under annual tillage and sub-days intensity changes
18 (Mullan et al. 2012). Again, it shows how crucial land management is for addressing soil erosion, and
19 the important role of rainfall intensity changes.

20 There is a large body of literature based on modelling future land degradation due to soil erosion
21 concluding that in spite of the increasing trend of erosive power of rainfall (*medium evidence, high*
22 *agreement*) land degradation is primarily determined by land management (*very high confidence*).

23 **4.4.1.2 Climate induced vegetation changes, implications for land degradation**

24 The spatial mosaic of vegetation is determined by three factors: the ability of species to reach a
25 particular location, how species tolerate the environmental conditions at that location (e.g.
26 temperature, precipitation, wind, the topographic and soil conditions), and the interaction between
27 species (including above/below ground species (Settele et al. 2015). Climate change is projected to
28 alter the conditions and hence impact the spatial mosaic of vegetation, which can be considered a
29 form of land degradation. Warren et al. (2018) estimated that only about 33% of globally important
30 biodiversity conservation areas will remain intact if global mean temperature increases to 4.5°C, while
31 twice that area (67%) will remain intact if warming is restricted to 2°C. According to AR5, the
32 clearest link between climate change and ecosystem change is when temperature is the primary driver,
33 with changes of Arctic tundra as a response to significant warming as the best example (Settele et al.
34 2015). Even though distinguishing climate induced changes from land use changes is challenging,
35 Boit et al. (2016) suggest that 5-6% of biomes in South America will undergo biome shifts until 2100,
36 regardless of scenario, attributed to climate change. The projected biome shifts are primarily forests
37 shifting to shrubland and dry forests becoming fragmented and isolated from more humid forests
38 (Boit et al. 2016). Boreal forests are subject to unprecedented warming in terms of speed and
39 amplitude (IPCC 2013b), with significant impacts on their regional distribution (Juday et al. 2015).
40 Globally, tree lines are generally expanding northward and to higher elevations, or remaining stable,
41 while a reduction in tree line was rarely observed and only where disturbances occurred (Harsch et al.
42 2009) There is *limited evidence* of a slow northward migration of the boreal forest in eastern North
43 America (Gamache and Payette 2005). The thawing of permafrost may increase drought induced tree
44 mortality throughout the circumboreal zone (Gauthier et al. 2015).

45 Forests are a prime regulator of hydrological cycling, both fluxes of atmospheric moisture and
46 precipitation, hence climate and forests are inextricably linked (Ellison et al. 2017; Keys et al. 2017).
47 Forest management influences the storage and flow of water in forested watersheds, particularly
48 harvesting, thinning and construction of roads increase the likelihood of floods as an outcome of

1 extreme climate events (Eisenbies et al. 2007). Water balance of at least partly forested landscapes is
2 to a large extent controlled by forest ecosystems (Sheil and Murdiyarso 2009; Pokam et al. 2014).
3 This includes surface runoff, as determined by evaporation and transpiration and soil conditions, and
4 water flow routing (Eisenbies et al. 2007). Water use efficiency (i.e., the ratio of water loss to biomass
5 gain) is increasing with increased CO₂ levels (Keenan et al. 2013), hence transpiration is predicted to
6 decrease which in turn will increase surface runoff (Schlesinger and Jasechko 2014). However, the
7 interaction of several processes makes predictions challenging (Frank et al. 2015; Trahan and
8 Schubert 2016). Surface runoff is an important agent in soil erosion.

9 Generally, removal of trees through harvesting or forest death (Anderegg et al. 2012) will reduce
10 transpiration and hence increase the runoff during the growing season. Management induced soil
11 disturbance (such as skid trails and roads) will affect water flow routing to rivers and streams (Zhang
12 et al. 2017; Luo et al. 2018; Eisenbies et al. 2007).

13 Climate change affects forests in both positive and negative ways (Trumbore et al. 2015; Price et al.
14 2013) and there will be regional and temporal differences in vegetation responses (Hember et al.
15 2017; Midgley and Bond 2015). Several climate change related drivers interact in complex ways, such
16 as warming, changes in precipitation and water balance, CO₂ fertilisation, and nutrient cycling, which
17 makes projections of future net impacts challenging (see 2.3.1.2) (Kurz et al. 2013; Price et al. 2013).
18 In high latitudes, a warmer climate will extend the growing seasons which however, could be
19 constrained by summer drought (Holmberg et al. 2019) while increasing levels of atmospheric CO₂
20 will increase water use efficiency but not necessarily tree growth (Giguère-Croteau et al. 2019).
21 Improving one growth limiting factor will only enhance tree growth if other factors are not limiting
22 (Norby et al. 2010; Trahan and Schubert 2016; Xie et al. 2016; Frank et al. 2015). Increasing forest
23 productivity has been observed in most of Fennoscandia (Kauppi et al. 2014; Henttonen et al. 2017),
24 Siberia and the northern reaches of North America as a response to a warming trend (Gauthier et al.
25 2015) but increased warming may also decrease forest productivity and increase risk of tree mortality
26 and natural disturbances (Price et al. 2013; Girardin et al. 2016; Beck et al. 2011; Hember et al. 2016;
27 Allen et al. 2011). The climatic conditions in high latitudes are changing at a magnitude faster than
28 the ability of forests to adapt with detrimental, yet unpredictable, consequences (Gauthier et al. 2015).

29 Negative impacts dominate, however, and have already been documented (Lewis et al. 2004; Bonan et
30 al. 2008; Beck et al. 2011) and are predicted to increase (Miles et al. 2004 ; Allen et al. 2010;
31 Gauthier et al. 2015; Girardin et al. 2016; Trumbore et al. 2015). Several authors have emphasized a
32 concern that tree mortality (forest dieback) will increase due to climate induced physiological stress as
33 well as interactions between physiological stress and other stressors, such as insect pests, diseases,
34 and wildfires (Anderegg et al. 2012; Sturrock et al. 2011; Bentz et al. 2010; McDowell et al. 2011).
35 Extreme events such as extreme heat and drought, storms, and floods also pose increased threats to
36 forests in both high and low latitude forests (Lindner et al. 2010; Mokria et al. 2015). However,
37 comparing observed forest dieback with modelled climate induced damages did not show a general
38 link between climate change and forest dieback (Steinkamp and Hickler 2015). Forests are subject to
39 increasing frequency and intensity of wildfires which is projected to increase substantially with
40 continued climate change (see also Cross-Chapter Box 3: Fire and climate change, Chapter 2) (Price
41 et al. 2013). In the tropics, interaction between climate change, CO₂ and fire could lead to abrupt
42 shifts between woodland and grassland dominated states in the future (Shanahan et al. 2016).

43 Within the tropics, much research has been devoted to understanding how climate change may alter
44 regional suitability of various crops. For example coffee is expected to be highly sensitive to both
45 temperature and precipitation changes, both in terms of growth and yield and in terms of increasing
46 problems of pests (Ovalle-Rivera et al. 2015). Some studies conclude that the global area of coffee
47 production will decrease by 50% (Bunn et al. 2015). Due to increased heat stress, the suitability of
48 Arabica coffee is expected to deteriorate in Mesoamerica, while it can improve in high altitude areas

1 in South America. The general pattern is that the climatic suitability for Arabica coffee will
2 deteriorate at low altitudes of the tropics as well as at the higher latitudes (Ovalle-Rivera et al. 2015).
3 This means that climate change in and of itself can render unsustainable previously sustainable land
4 use and land management practices and vice versa (Laderach et al. 2011).

5 Rangelands are projected to change in complex ways due to climate change. Increasing levels of
6 atmospheric CO₂ stimulate directly plant growth and can potentially compensate negative effects from
7 drying by increasing rain use efficiency. But the positive effect of increasing CO₂ will be mediated by
8 other environmental conditions, primarily water availability but also nutrient cycling, fire regimes and
9 invasive species. Studies over the North American rangelands suggest, for example, that warmer and
10 dryer climatic conditions will reduce NPP in the southern Great Plains, the Southwest, and northern
11 Mexico, but warmer and wetter conditions will increase NPP in the northern Plains and southern
12 Canada (Polley et al. 2013).

13 **4.4.1.3 Coastal erosion**

14 Coastal erosion is expected to increase dramatically by sea level rise and in some areas in
15 combination with increasing intensity of cyclones (highlighted in Section 4.9.6). Cyclone induced
16 coastal erosion). Coastal regions are also characterised by high population density, particularly in Asia
17 (Bangladesh, China, India, Indonesia, Vietnam), whereas the highest population increase in coastal
18 regions is projected in Africa (East Africa, Egypt, and West Africa) (Neumann et al. 2015). For
19 coastal regions worldwide, and particularly in developing countries with high population density in
20 low-lying coastal areas, limiting the warming to 1.5°C to 2.0 °C will have major socio-economic
21 benefits compared with higher temperature scenarios (IPCC 2018a; Nicholls et al. 2018). For more in-
22 depth discussions on coastal process, please refer to Chapter 4 of the upcoming IPCC Special Report
23 on The Ocean and Cryosphere in a Changing Climate (IPCC SROCC).

24 Despite the uncertainty related to the responses of the large ice sheets of Greenland and west
25 Antarctica, climate change-induced sea level rise is largely accepted and represents one of the biggest
26 threats faced by coastal communities and ecosystems (Nicholls et al. 2011; Cazenave and Cozannet
27 2014; DeConto and Pollard 2016; Mengel et al. 2016). With significant socio-economic effects, the
28 physical impacts of projected sea level rise, notably coastal erosion, have received considerable
29 scientific attention (Nicholls et al. 2011; Rahmstorf 2010; Hauer et al. 2016).

30 Rates of coastal erosion or recession will increase due to rising sea levels and in some regions also in
31 combination with increasing oceans waves (Day and Hodges 2018; Thomson and Rogers 2014;
32 McInnes et al. 2011; Mori et al. 2010), lack or absence of sea-ice (Savard et al. 2009; Thomson and
33 Rogers 2014) and thawing of permafrost (Hoegh-Guldberg et al. 2018), and changing cyclone paths
34 (Tamarin-Brodsky and Kaspi 2017; Lin and Emanuel 2016a). The respective role of the different
35 climate factors in the coastal erosion process will vary spatially. Some studies have shown that the
36 role of sea level rise on the coastal erosion process can be less important than other climate factors,
37 like wave heights, changes in the frequency of the storms, and the cryogenic processes (Ruggiero
38 2013; Savard et al. 2009). Therefore, in order to have a complete picture of the potential effects of sea
39 level rise on rates of coastal erosion, it is crucial to consider the combined effects of the
40 aforementioned climate controls and the geomorphology of the coast under study.

41 Coastal wetlands around the world are sensitive to sea-level rise. Projections of the impacts on global
42 coastlines are inconclusive, with some projections suggesting that 20% to 90% (depending on sea-
43 level rise scenario) of present day wetlands will disappear during the 21st century (Spencer et al.
44 2016). Another study, which included natural feed-back processes and management responses
45 suggested that coastal wetlands may actually increase (Schuerch et al. 2018b).

1 Low-lying coastal areas in the tropics are particularly subject to the combined effect of sea-level rise
2 and increasing intensity of tropical cyclones, conditions which in many cases pose limits to
3 adaptation, see section 4.8.5.1.

4 Many large coastal deltas are subject to the additional stress of shrinking deltas as a consequence of
5 the combined effect of reduced sediment loads from rivers due to damming and water use, and land
6 subsidence resulting from extraction of ground water or natural gas, and aquaculture (Higgins et al.
7 2013; Tessler et al. 2016; Minderhoud et al. 2017; Tessler et al. 2015; Brown and Nicholls 2015;
8 Szabo et al. 2016; Yang et al. 2019; Shirzaei and Bürgmann 2018; Wang et al. 2018; Fuangswasdi et
9 al. 2019). In some cases the rate of subsidence can outpace the rate of sea level rise by one order of
10 magnitude (Minderhoud et al. 2017) or even two (Higgins et al. 2013). Recent findings from the
11 Mississippi Delta raises the risk of a systematic underestimation of the rate of land subsidence in
12 coastal deltas (Keogh and Törnqvist 2019)

13 In sum, from a land degradation point of view, low lying coastal areas are particularly exposed to the
14 nexus of climate change and increasing concentration of people (Elliott et al. 2014) (*robust evidence,*
15 *high agreement*) and the situation will become particularly acute in delta areas shrinking from both
16 reduced sediment loads and land subsidence (*robust evidence, high agreement*).

17 **4.4.2 Indirect impacts on land degradation**

18 Indirect impacts of climate change on land degradation are difficult to quantify because of the many
19 conflating factors. The causes of land-use change are complex, combining physical, biological and
20 socioeconomic drivers (Lambin et al. 2001; Lambin and Meyfroidt 2011). One such driver of land-use
21 change is the degradation of agricultural land, which can result in a negative cycle of natural land
22 being converted to agricultural land to sustain production levels. The intensive management of
23 agricultural land can lead to a loss of soil function, negatively impacting the many ecosystem services
24 provided by soils including maintenance of water quality and soil carbon sequestration (Smith et al.
25 2016a). The degradation of soil quality due to cropping is of particular concern in tropical regions,
26 where it results in a loss of productive potential of the land, affecting regional food security and
27 driving conversion of non-agricultural land, such as forestry, to agriculture (Lambin et al. 2003;
28 Drescher et al. 2016; Van der Laan et al. 2017). Climate change will exacerbate these negative cycles
29 unless sustainable land managed practices are implemented.

30 Climate change impacts on agricultural productivity (see Chapter 5) will have implications for the
31 intensity of land use and hence exacerbate the risk of increasing land degradation. There will be both
32 localised effects (i.e., climate change impacts on productivity affecting land use in the same region)
33 and teleconnections (i.e., climate change impacts and land-use change are spatially and temporally
34 separate) (Wicke et al. 2012; Pielke et al. 2007). If global temperature increases beyond 3°C it will
35 have negative yield impacts on all crops (Porter et al. 2014) which, in combination with a doubling of
36 demands by 2050 (Tilman et al. 2011), and increasing competition for land from the expansion of
37 negative emissions technologies (IPCC 2018a; Schleussner et al. 2016), will exert strong pressure on
38 agricultural lands and food security.

39 In sum, reduced productivity of most agricultural crops will drive land-use changes worldwide (*robust*
40 *evidence, medium agreement*), but predictions of how this will impact land degradation is challenging
41 because of several conflating factors. Social change, such as widespread changes in dietary
42 preferences will have a huge impact on agriculture and hence land degradation (*medium evidence,*
43 *high agreement*).

4.5 Impacts of bioenergy and technologies for CO₂ removal (CDR) on land degradation

4.5.1 Potential scale of bioenergy and land-based CDR

In addition to the traditional land use drivers (e.g. population growth, agricultural expansion, forest management), a new driver will interact to increase competition for land throughout this century: the potential large-scale implementation of land-based technologies for CO₂ removal (CDR). Land-based CDR include afforestation and reforestation, bioenergy with carbon capture and storage (BECCS), soil carbon management, biochar and enhanced weathering (Smith et al., 2015; Smith 2016)

Most scenarios, including two of the four pathways in the IPCC Special Report on 1.5°C (IPCC 2018a), compatible with stabilisation at 2°C involve substantial areas devoted to land-based CDR, specifically afforestation/ reforestation and BECCS (Schleussner et al. 2016; Smith et al. 2016b; Mander et al. 2017). Even larger land areas are required in most scenarios aimed at keeping average global temperature increases to below 1.5 °C, and scenarios that avoid BECCS also require large areas of energy crops in many cases (IPCC 2018b), although some options with strict demand-side management avoid this need (Grubler et al. 2018). Consequently, the addition of carbon capture and storage (CCS) systems to bioenergy facilities enhances mitigation benefits because it increases the carbon retention time and reduces emissions relative to bioenergy facilities without CCS. The IPCC SR15 states that “When considering pathways limiting warming to 1.5°C with no or limited overshoot, the full set of scenarios shows a conversion of 0.5 – 11 M km² of pasture into 0 – 6 M km² for energy crops, a 2 M km² reduction to 9.5 M km² increase forest, and a 4 M km² decrease to a 2.5 M km² increase in non-pasture agricultural land for food and feed crops by 2050 relative to 2010.” (Rogelj et al., 2018, p. 145). For comparison, the global cropland area in 2010 was 15.9 M km² (Table 1.1), and (Woods et al. 2015) estimate the area of abandoned and degraded land potentially available for energy crops (or afforestation/reforestation) exceeds 5 M km². However, the area of available land has long been debated, as much marginal land is subject customary land tenure and used informally often by impoverished communities (Baka 2013, 2014; Haberl et al. 2013; Young 1999). Thus, as noted in the SR15, “The implementation of land-based mitigation options would require overcoming socio-economic, institutional, technological, financing and environmental barriers that differ across regions” (IPCC, 2018a, p. 18).

The wide range of estimates reflects the large differences among the pathways, availability of land in various productivity classes, types of NET implemented, uncertainties in computer models, and social and economic barriers to implementation (Fuss et al. 2018; Nemet et al. 2018; Minx et al. 2018).

4.5.2 Risks of land degradation from expansion of bioenergy and land-based CDR

The large-scale implementation of high intensity dedicated energy crops, and harvest of crop and forest residues for bioenergy, could contribute to increases in the area of degraded lands: intensive land management can result in nutrient depletion, over fertilisation and soil acidification, salinisation (from irrigation without adequate drainage), wet ecosystems drying (from increased evapotranspiration), as well as novel erosion and compaction processes (from high impact biomass harvesting disturbances) and other land degradation processes described in Section 4.2.1.

Global integrated assessment models used in the analyses of mitigation pathways vary in their approaches to modelling CDR (Bauer et al. 2018) and the outputs have large uncertainties due to their limited capability to consider site-specific details (Krause et al. 2018). Spatial resolutions vary from 11 world regions to 0.25 degrees gridcells (Bauer et al. 2018). While model projections identify potential areas for CDR implementation (Heck et al. 2018), the interaction with climate change induced biome shifts, available land and its vulnerability to degradation are unknown. The crop/forest

1 types and management practices that will be implemented are also unknown, and will be influenced
2 by local incentives and regulations. While it is therefore currently not possible to project the area at
3 risk of degradation from the implementation of land-based CDR, there is a clear risk that expansion of
4 energy crops at the scale anticipated could put significant strain on land systems, biosphere integrity,
5 freshwater supply and biogeochemical flows (Heck et al. 2018). Similarly, extraction of biomass for
6 energy from existing forests, particularly where stumps are utilized, can impact soil health (de Jong et
7 al. 2017). Reforestation and afforestation present a lower risk of land degradation and may in fact
8 reverse degradation (see Section 4.5.3) although potential adverse hydrological and biodiversity
9 impacts will need to be managed (Caldwell et al. 2018; Brinkman et al. 2017). Soil carbon
10 management can deliver negative emissions while reducing or reversing land degradation. Chapter 6
11 discusses the significance of context and management in determining environmental impacts of
12 implementation of land-based options.

13 **4.5.3 Potential contributions of land-based CDR to reducing and reversing land** 14 **degradation**

15 Although large-scale implementation of land-based CDR has significant potential risks, the need for
16 negative emissions and the anticipated investments to implement such technologies can also create
17 significant opportunities. Investments into land-based CDR can contribute to halting and reversing
18 land degradation, to the restoration or rehabilitation of degraded and marginal lands (Chazdon and
19 Uriarte 2016; Fritsche et al. 2017) and can contribute to the goals of land degradation neutrality (Orr
20 et al. 2017a).

21 Estimates of the global area of degraded land range from less than 10 to 60 M km² (Gibbs and
22 Salmon 2015), see also section 4.3.1. Additionally, large areas are classified as marginal lands and
23 may be suitable for the implementation of bioenergy and land-based CDR (Woods et al. 2015). The
24 yield per hectare of marginal and degraded lands is lower than on fertile lands, and if CDR will be
25 implemented on marginal and degraded lands this will increase the area demand and costs per unit
26 area of achieving negative emissions (Fritsche et al. 2017). Selection of lands suitable for CDR must
27 be considered carefully to reduce conflicts with existing users, to assess the possible trade-offs in
28 biodiversity contributions of the original and the CDR land uses, to quantify the impacts on water
29 budgets, and to ensure sustainability of the CDR land use.

30 Land use and land condition prior to the implementation of CDR affect the climate change benefits
31 (Harper et al. 2018). Afforestation/ reforestation on degraded lands can increase C stocks in
32 vegetation and soil, increase carbon sinks (Amichev et al. 2012), and deliver co-benefits for
33 biodiversity and ecosystem services particularly if a diversity of local species are used. Afforestation
34 and reforestation on native grasslands can reduce soil carbon stocks, although the loss is typically
35 more than compensated by increases in biomass and dead organic matter C stocks (Bárcena et al.
36 2014; Li et al. 2012; Ovalle-Rivera et al. 2015; Shi et al. 2013), and may impact biodiversity (Li et al.
37 2012) (see also 4.4.1: Large scale forest cover expansion, what can be learned in context of the
38 SRCCL).

39 Strategic incorporation of energy crops into agricultural production systems, applying an integrated
40 landscape management approach, can provide co-benefits for management of land degradation and
41 other environmental objectives. For example, buffers of Miscanthus and other grasses can enhance
42 soil carbon and reduce water pollution (Cacho et al. 2018; Odgaard et al. 2019), and strip-planting of
43 short rotation tree crops can reduce the water table where crops are affected by dryland salinity
44 (Robinson et al. 2006). Shifting to perennial grain crops has the potential to combine food production
45 with carbon sequestration at a higher rate than with annual grain crops and avoid the trade-off
46 between food production and climate change mitigation (Crews, Carton, & Olsson, 2018; de Oliveira,
47 Brunsell, Sutherlin, Crews, & DeHaan, 2018; Ryan et al., 2018, see also 4.9.2).

1 Changes in land cover can affect surface reflectance, water balances and emissions of volatile organic
2 compounds and thus the non-GHG impacts on the climate system from afforestation/reforestation or
3 planting energy crops (Anderson et al. 2011; Bala et al. 2007; Betts 2000; Betts et al. 2007), (see
4 Section 4.6 for further details). Some of these impacts reinforce the GHG mitigation benefits, while
5 others off-set the benefits, with strong local (slope, aspect) and regional (boreal vs. tropical biomes)
6 differences in the outcomes (Li et al. 2015). Adverse effects on albedo from afforestation with
7 evergreen conifers in boreal zones can be reduced through planting of broadleaf deciduous species
8 (Astrup et al. 2018; Cai et al. 2011a; Anderson et al. 2011).

9 Combining CDR technologies may prove synergistic. Two soil management techniques with an
10 explicit focus on increasing the soil carbon content rather than promoting soil conservation more
11 broadly have been suggested: Addition of biochar to agricultural soils (see 4.9.5) and addition of
12 ground silicate minerals to soils in order to take up atmospheric CO₂ through chemical weathering
13 (Taylor et al. 2017; Haque et al. 2019; Beerling 2017; Strefler et al. 2018). The addition of biochar is
14 comparatively well understood and also field tested at large scale, see section 4.9.5 for a
15 comprehensive discussion. The addition of silicate minerals to soils is still highly uncertain in terms
16 of its potential (from 95 GtCO₂ yr⁻¹ (Strefler et al. 2018) to only 2-4 GtCO₂ yr⁻¹ (Fuss et al. 2018)) and
17 costs (Schlesinger and Amundson 2018).

18 Effectively addressing land degradation through implementation of bioenergy and land-based CDR
19 will require site-specific local knowledge, matching of species with the local land, water balance,
20 nutrient and climatic conditions, and ongoing monitoring and, where necessary, adaptation of land
21 management to ensure sustainability under global change (Fritsche et al. 2017). Effective land
22 governance mechanisms including integrated land-use planning, along with strong sustainability
23 standards could support deployment of energy crops and afforestation/reforestation at appropriate
24 scales and geographical contexts (Fritsche et al. 2017). Capacity-building and technology transfer
25 through the international cooperation mechanisms of the Paris Agreement could support such efforts.
26 Modelling to inform policy development is most useful when undertaken with close interaction
27 between model developers and other stakeholders including policymakers to ensure that models
28 account for real world constraints (Dooley and Kartha 2018).

29 International initiatives to restore lands, such as the Bonn Challenge (Verdone and Seidl 2017) and
30 the New York Declaration on Forests (Chazdon et al. 2017), and interventions undertaken for Land
31 Degradation Neutrality and implementation of NDCs (see Glossary) can contribute to NET objectives.
32 Such synergies may increase the financial resources available to meet multiple objectives (see section
33 4.8.4).

34 **4.5.4 Traditional biomass provision and land degradation**

35 Traditional biomass (fuelwood, charcoal, agricultural residues, animal dung) used for cooking and
36 heating by some 2.8 billion people (38% of global population) in non-OECD countries accounts for
37 more than half of all bioenergy used worldwide (IEA 2017; REN21 2018; see Cross-Chapter Box 7
38 on Bioenergy, Chapter 6). Cooking with traditional biomass has multiple negative impacts on human
39 health, particularly for women, children and youth (Machisa et al. 2013; Sinha and Ray 2015; Price
40 2017; Mendum and Njenga 2018; Adefuye et al. 2007) and on household productivity including high
41 workloads for women and youth (Mendum and Njenga 2018; Brunner et al. 2018; Hou et al. 2018;
42 Njenga et al. 2019). Traditional biomass is land-intensive due to reliance on open fires, inefficient
43 stoves and overharvesting of woodfuel, contributing to land degradation, losses in biodiversity and
44 reduced ecosystem services (IEA 2017; Bailis et al. 2015; Masera et al. 2015; Specht et al. 2015;
45 Fritsche et al. 2017; Fuso Nerini et al. 2017). Traditional woodfuels account for 1.9-2.3% of global
46 GHG emissions, particularly in “hotspots” of land degradation and fuelwood depletion in eastern
47 Africa and South Asia, such that one-third of traditional woodfuels globally are harvested

1 unsustainably (Bailis et al. 2015). Scenarios to significantly reduce reliance on traditional biomass in
2 developing countries present multiple co-benefits (*high evidence, high agreement*), including reduced
3 emissions of black carbon, a short-lived climate forcer that also causes respiratory disease (Shindell et
4 al. 2012).

5 A shift from traditional to modern bioenergy, especially in the African context, contributes to
6 improved livelihoods and can reduce land degradation and impacts on ecosystem services (Smeets et
7 al. 2012; Gasparatos et al. 2018; Mudombi et al. 2018). In Sub-Saharan Africa, most countries
8 mention woodfuel in their Nationally Determined Contribution (NDC) but fail to identify
9 transformational processes to make fuelwood a sustainable energy source compatible with improved
10 forest management (Amugune et al. 2017). In some regions, especially in South and Southeast Asia, a
11 scarcity of woody biomass may lead to excessive removal and use of agricultural wastes and residues,
12 which contributes to poor soil quality and land degradation (Blanco-Canqui and Lal 2009; Mateos et
13 al. 2017).

14 In sub-Saharan Africa, forest degradation is widely associated with charcoal production although in
15 some tropical areas rapid re-growth can offset forest losses (Hoffmann et al. 2017; McNicol et al.
16 2018). Overharvesting of wood for charcoal contributes to the high rate of deforestation in sub-
17 Saharan Africa, which is five times the world average, due in part to corruption and weak governance
18 systems (Sulaiman et al. 2017). Charcoal may also be a by-product of forest clearing for agriculture,
19 with charcoal sale providing immediate income when the land is cleared for food crops (Kiruki et al.
20 2017; Ndegwa et al. 2016). Besides loss of forest carbon stock, a further concern for climate change is
21 methane and black carbon emissions from fuelwood burning and traditional charcoal-making
22 processes (Bond et al. 2013; Patange et al. 2015; Sparrevik et al. 2015).

23 A fundamental difficulty in reducing environmental impacts associated with charcoal lies in the small-
24 scale nature of much charcoal production in sub-Saharan Africa leading to challenges in regulating its
25 production and trade, which is often informal, and in some cases illegal, but nevertheless widespread
26 since charcoal is the most important urban cooking fuel (Zulu 2010; Zulu and Richardson 2013; Smith
27 et al. 2015; World Bank 2009) (World Bank, 2009). Urbanisation combined with population growth
28 has led to continuously increasing charcoal production. Low efficiency of traditional charcoal
29 production results in a four-fold increase in raw woody biomass required and thus much greater
30 biomass harvest (Hojas-Gascon et al. 2016; Smeets et al. 2012). With continuing urbanisation
31 anticipated, increased charcoal production and use will probably contribute to increasing land
32 pressures and increased land degradation, especially in sub-Saharan Africa (*medium evidence, high
33 agreement*).

34 Although it could be possible to source this biomass more sustainably, the ecosystem and health
35 impacts of this increased demand for cooking fuel would be reduced through use of other renewable
36 fuels or in some cases, non-renewable fuels (LPG), as well as through improved efficiency in end-use
37 and through better resource and supply chain management (Santos et al. 2017; Smeets et al. 2012;
38 Hoffmann et al. 2017). Integrated response options such as agro-forestry (see Chapter 6) and good
39 governance mechanisms for forest and agricultural management (see Chapter 7) can support the
40 transition to sustainable energy for households and reduce the environmental impacts of traditional
41 biomass.

42 **4.6 Impacts of land degradation on climate**

43 While Chapter 2 has its focus on land cover changes and their impacts on the climate system, this
44 chapter focuses on the influences of individual land degradation processes on climate (see cross
45 chapter Table 4.1) which may or may not take place in association to land cover changes. The effects

1 of land degradation on CO₂ and other greenhouse gases as well as those on surface albedo and other
2 physical controls of the global radiative balance are discussed.

3 **4.6.1 Impacts on greenhouse gases**

4 Land degradation processes with direct impact on soil and terrestrial biota have great relevance in
5 terms of CO₂ exchange with the atmosphere given the magnitude and activity of these reservoirs in
6 the global C cycle. As the most widespread form of soil degradation, erosion detaches the surface soil
7 material which typically hosts the highest organic C stocks, favoring the mineralisation and release as
8 CO₂, yet complementary processes such as C burial may compensate this effect, making soil erosion a
9 long-term C sink (*low agreement, limited evidence*), (Wang et al. 2017b), but see also (Chappell et al.
10 2016). Precise estimation of the CO₂ released from eroded lands is challenged by the fact that only a
11 fraction of the detached C is eventually lost to the atmosphere. It is important to acknowledge that a
12 substantial fraction of the eroded material may preserve its organic C load in field conditions.
13 Moreover, C sequestration may be favored through the burial of both the deposited material and the
14 surface of its hosting soil at the deposition location (Quinton et al. 2010). The cascading effects of
15 erosion on other environmental processes at the affected sites can often cause net CO₂ emissions
16 through their indirect influence on soil fertility and the balance of organic C inputs and outputs,
17 interacting with other non-erosive soil degradation processes such as nutrient depletion, compaction
18 and salinisation, which can lead to the same net C effects (see Table 4.1) (van de Koppel et al. 1997).

19 As natural and human-induced erosion can result in net C storage in very stable buried pools at the
20 deposition locations, degradation in those locations has a high C-release potential. Coastal ecosystems
21 such as mangrove forests, marshes and seagrasses are a typical deposition locations and their
22 degradation or replacement with other vegetation is resulting in a substantial C release (0.15 to 1.02
23 Gt C yr⁻¹) (Pendleton et al. 2012), which highlights the need for a spatially-integrated assessment of
24 land degradation impacts on climate that considers in-situ but also ex-situ emissions.

25 Cultivation and agricultural management of cultivated land are relevant in terms of global CO₂ land-
26 atmosphere exchange (see also 4.8.1). Besides the initial pulse of CO₂ emissions associated with the
27 onset of cultivation and associated vegetation clearing (see Chapter 2), agricultural management
28 practices can increase or reduce C losses to the atmosphere. Although global croplands are considered
29 to be at relatively neutral stage in the current decade (Houghton et al. 2012), this results from a highly
30 uncertain balance between coexisting net losses and gains. Degradation losses of soil and biomass
31 carbon appear to be compensated by gains from soil protection and restoration practices such as cover
32 crops, conservation tillage and nutrient replenishment favoring organic matter build-up. Cover crops,
33 increasingly used to improve soils, have the potential to sequester 0.12 Gt C yr⁻¹ on global croplands
34 with a saturation time of more than 150 years (Poeplau and Don 2015). No-till practices (i.e. tillage
35 elimination favoring crop residue retention in the soil surface) which were implemented to protect
36 soils from erosion and reduce land preparation times, were also seen with optimism as a C
37 sequestration option, which today is considered more modest globally and, in some systems, even less
38 certain (VandenBygaart 2016; Cheesman et al. 2016; Powlson et al. 2014). Among soil fertility
39 restoration practices, lime application for acidity correction, increasingly important in tropical
40 regions, can generate a significant net CO₂ source in some soils (Bernoux et al. 2003, Alemu et al
41 2017).

42 Land degradation processes in seminatural ecosystems driven by unsustainable uses of their
43 vegetation through logging or grazing lead to reduced plant cover and biomass stocks, causing net C
44 releases from soils and plant stocks. Degradation by logging activities is particularly important in
45 developing tropical and subtropical regions, involving C releases that exceed by far the biomass of
46 harvested products, including additional vegetation and soil sources that are estimated to reach 0.6 Gt
47 C yr⁻¹ (Pearson et al. 2014, 2017). Excessive grazing pressures pose a more complex picture with

1 variable magnitudes and even signs of C exchanges. A general trend of higher C losses in humid
2 overgrazed rangelands suggests a high potential for C sequestration following the rehabilitation of
3 those systems (Conant and Paustian 2002) with a global potential sequestration of 0.045 Gt C yr⁻¹. A
4 special case of degradation in rangelands are those processes leading to the woody encroachment of
5 grass-dominated systems, which can be responsible of declining animal production but high C
6 sequestration rates (Asner et al. 2003, Maestre et al. 2009).

7 Fire regime shifts in wild and seminatural ecosystems can become a degradation process in itself, with
8 high impact on net C emission and with underlying interactive human and natural drivers such as
9 burning policies (Van Wilgen et al. 2004), biological invasions (Brooks et al. 2009), and plant
10 pest/disease spread (Kulakowski et al. 2003). Some of these interactive processes affecting
11 unmanaged forests have resulted in massive C release, highlighting how degradation feedbacks on
12 climate are not restricted to intensively used land but can affect wild ecosystems as well (Kurz et al.
13 2008).

14 Agricultural land and wetlands represent the dominant source of non-CO₂ greenhouse gases (Chen et
15 al. 2018d). In agricultural land, the expansion of rice cultivation (increasing CH₄ sources), ruminant
16 stocks and manure disposal (increasing CH₄, N₂O and NH₃ fluxes) and nitrogen over-fertilisation
17 combined with soil acidification (increasing N₂O fluxes) are introducing the major impacts (*medium*
18 *agreement, medium evidence*) and their associated emissions appear to be exacerbated by global
19 warming (*medium agreement and medium evidence*) (Oertel et al. 2016).

20 As the major sources of global N₂O emissions, over-fertilisation and manure disposal are not only
21 increasing in-situ sources but also stimulating those along the pathway of dissolved inorganic nitrogen
22 transport all the way from draining waters to the ocean (*high agreement, medium evidence*). Current
23 budgets of anthropogenically fixed nitrogen on the Earth System (Tian et al. 2015; Schaefer et al.
24 2016; Wang et al. 2017a) suggest that N₂O release from terrestrial soils and wetlands accounts for 10-
25 15% of the inputs, yet many further release fluxes along the hydrological pathway remain uncertain,
26 with emissions from oceanic “dead-zones” being a major aspect of concern (Schlesinger 2009;
27 Rabalais et al. 2014).

28 Environmental degradation processes focused on the hydrological system, which are typically
29 manifested at the landscape scale, include both drying (as in drained wetlands or lowlands) and
30 wetting trends (as in waterlogged and flooded plains). Drying of wetlands reduces CH₄ emissions
31 (Turetsky et al. 2014) but favors pulses of organic matter mineralization linked to high N₂O release
32 (Morse and Bernhardt 2013; Norton et al. 2011). The net warming balance of these two effects is not
33 resolved and may be strongly variable across different types of wetlands. In the case of flooding of
34 non-wetland soils, a suppression of CO₂ release is typically over compensated in terms of net
35 greenhouse impact by enhanced CH₄ fluxes, that stem from the lack of aeration but are aided by the
36 direct effect of extreme wetting on the solubilisation and transport of organic substrates (McNicol and
37 Silver 2014). Both wetlands rewetting/restoration and artificial creation can increase CH₄ release
38 (Altor and Mitsch 2006; Fenner et al. 2011). Permafrost thawing is another major source of CH₄
39 release with substantial long-term contributions to the atmosphere that are starting to get globally
40 quantified (Christensen et al. 2004; Schuur et al. 2015; Walter Anthony et al. 2016).

41 **4.6.2 Physical impacts**

42 Among the physical effects of land degradation, surface albedo changes are those with the most
43 evident impact on the net global radiative balance and net climate warming/cooling. Degradation
44 processes affecting wild and semi-natural ecosystems such as fire regime changes, woody
45 encroachment, logging and overgrazing can trigger strong albedo changes before significant
46 biogeochemical shifts take place, in most cases these two types of effects have opposite signs in terms

1 of net radiative forcing, making their joint assessment critical for understanding climate feedbacks
2 (Bright et al. 2015).

3 In the case of forest degradation or deforestation, the albedo impacts are highly dependent on the
4 latitudinal/climatic belt to which they belong. In boreal forests the removal or degradation of the tree
5 cover increases albedo (net cooling effect)(*medium evidence, high agreement*) as the reflective snow
6 cover becomes exposed, which can exceed the net radiative effect of the associated C release to the
7 atmosphere (Davin et al. 2010; Pinty et al. 2011). On the other hand, progressive greening of boreal
8 and temperate forests has contributed to net albedo declines (*medium agreement, medium evidence*)
9 (Planque et al. 2017; Li et al. 2018a). In the northern treeless vegetation belt (tundra), shrub
10 encroachment leads to the opposite effect as the emergence of plant structures above the snow cover
11 level reduce winter-time albedo (Sturm 2005).

12 The extent to which albedo shifts can compensate carbon storage shifts at the global level has not
13 been estimated. A significant but partial compensation takes place in temperate and subtropical dry
14 ecosystems in which radiation levels are higher and C stocks smaller compared to their more humid
15 counterparts (*medium agreement, medium evidence*). In cleared dry woodlands half of the net global
16 warming effect of net C release has been compensated by albedo increase (Houspanossian et al.
17 2013), whereas in afforested dry rangelands albedo declines cancelled one fifth of the net C
18 sequestration (Rotenberg and Yakir 2010). Other important cases in which albedo effects impose a
19 partial compensation of C exchanges are the vegetation shifts associated to wild fires, as shown for
20 the savannahs, shrublands and grasslands of sub-Saharan Africa (Dintwe et al. 2017). Besides the net
21 global effects discussed above, albedo shifts can play a significant role on local climate (*high*
22 *agreement, medium evidence*), as exemplified by the effect of no-till agriculture reducing local heat
23 extremes in European landscapes (Davin et al. 2014) and the effects of woody encroachment causing
24 precipitation rises in the North American Great Plains (Ge and Zou 2013). Modeling efforts that
25 integrate ground data from deforested areas worldwide accounting for both physical and
26 biogeochemical effects, indicate that massive global deforestation would have a net warming impact
27 (Lawrence and Vandecar 2015) at both local and global levels with highlight non-linear effects of
28 forest loss on climate variables.

29 Beyond the albedo effects presented above, other physical impacts of land degradation on the
30 atmosphere can contribute to global and regional climate change. Of particular continental to global
31 relevance are the net cooling effects of dust emissions (*low agreement, medium evidence*) (Lau and
32 Kim 2007), but see also (Huang et al. 2014). Anthropogenic emission of mineral particles from
33 degrading land appear to have a similar radiative impact than all other anthropogenic aerosols
34 (Sokolik and Toon 1996). Dust emissions may explain regional climate anomalies through reinforcing
35 feedbacks, as suggested for the amplification of the intensity, extent and duration of the low
36 precipitation anomaly of the North American “Dust Bowl” in the 1930s (Cook et al. 2009). Another
37 source of physical effects on climate are surface roughness changes which, by affecting atmospheric
38 drag, can alter cloud formation and precipitation (*low agreement, low evidence*), as suggested by
39 modeling studies showing how the massive deployment of solar panels in the Sahara could increase
40 rainfall in the Sahel (Li et al. 2018c) or how woody encroachment in the Arctic tundra could reduce
41 cloudiness and raise temperature (Cho et al. 2018). The complex physical effects of deforestation, as
42 explored through modeling, converge into general net regional precipitation declines, tropical
43 temperature increases and boreal temperature declines, while net global effects are less certain
44 (Perugini et al. 2017). Integrating all the physical effects of land degradation and its recovery or
45 reversal is still challenge, yet modeling attempts suggest that over the last three decades the slow but
46 persistent net global greening caused by the average increase of leaf area in the land has caused a net
47 cooling of the Earth, mainly through the raise of evapotranspiration (Zeng et al. 2017) (*low*
48 *confidence*).

4.7 Impacts of climate-related land degradation on poverty and livelihoods

Unravelling the impacts of climate-related land degradation on poverty and livelihoods is highly challenging. This complexity is due to the interplay of multiple social, political, cultural, and economic factors, such as markets, technology, inequality, population growth, (Barbier and Hochard 2018) each of which interact and shape the ways in which social-ecological systems respond (Morton 2007). We find *limited evidence* attributing the impacts of climate-related land degradation to poverty and livelihoods, with climate often not distinguished from any other driver of land degradation. Climate is nevertheless frequently noted as a risk multiplier for both land degradation and poverty (*high agreement, robust evidence*) and is one of many stressors people live with, respond to and adapt to in their daily lives (Reid and Vogel 2006). Climate change is considered to exacerbate land degradation and potentially accelerate it due to heat stress, drought, changes to evapotranspiration rates and biodiversity, as well as a result of changes to environmental conditions that allow new pests and diseases to thrive (Reed and Stringer 2016b). In general terms, the climate (and climate change) can increase human and ecological communities' sensitivity to land degradation. Land degradation then leaves livelihoods more sensitive to the impacts of climate change and extreme climatic events (*high agreement, robust evidence*). If human and ecological communities exposed to climate change and land degradation are sensitive and cannot adapt, they can be considered vulnerable to it; if they are sensitive and can adapt, they can be considered resilient (Reed and Stringer 2016b). The impacts of land degradation will vary under a changing climate both spatially and temporally, leading some communities and ecosystems to be more vulnerable or more resilient than others under different scenarios. Even within communities, groups such as women and the youth are often more vulnerable than others.

4.7.1 Relationships between land degradation, climate change and poverty

This section sets out the relationships between land degradation and poverty, and climate change and poverty, leading to inferences about the 3-way links between them. Poverty is multidimensional and includes a lack of access to the whole range of capital assets that can be used to pursue a livelihood. Livelihoods constitute the capabilities, assets, and activities that are necessary to make a living (Chambers and Conway 1992; Olsson et al. 2014b).

The literature shows *high agreement* in terms of speculation that there are potential links between land degradation and poverty. However, studies have not provided robust quantitative assessments of the extent and incidence of poverty within land degradation affected populations (Barbier and Hochard 2016). Some researchers, e.g. Nachtergaele et al. (2011) estimate that 1.5 billion people were dependent upon degraded land to support their livelihoods in 2007, while >42 % of the world's poor population inhabit degraded areas. However, there is overall *low confidence* in the evidence base, a lack of studies that look beyond the past and present, and the literature calls for more in-depth research to be undertaken on these issues (Gerber et al. 2014). Recent work by Barbier and Hochard (Barbier and Hochard 2018) points to biophysical constraints such as poor soils and limited rainfall which interact to limit land productivity, suggesting that those farming in climatically less favourable agricultural areas are challenged by poverty. Studies such as those by (Coomes et al. 2011), focusing on an area in the Amazon, highlight the importance of the initial conditions of land holding in the dominant (shifting) cultivation system in terms of long-term effects on household poverty and future forest cover, showing initial land tenure and socio-economic aspects can make some areas less favourable too.

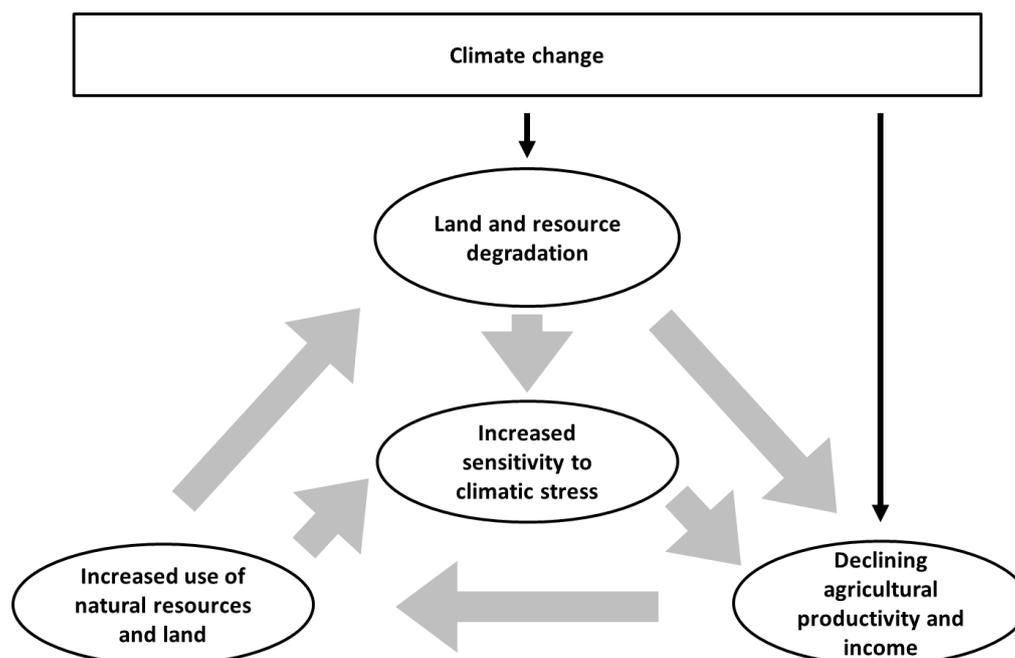
Much of the qualitative literature is focused on understanding the livelihood and poverty impacts of degradation through a focus on subsistence agriculture, where farms are small, under traditional or

1 informal tenure and where exposure to environmental (including climate) risks is high (Morton 2007).
2 In these situations, the poor lack access to assets (financial, social, human, natural and physical) and
3 in the absence of appropriate institutional supports and social protection, this leaves them sensitive
4 and unable to adapt, so a vicious cycle of poverty and degradation can ensue. To further illustrate the
5 complexity, livelihood assessments often focus on a single snapshot in time, livelihoods are dynamic
6 and people alter their livelihood activities and strategies depending the on internal and external
7 stressors to which they are responding (O'Brien et al. 2004). When certain livelihood activities and
8 strategies become no longer tenable as a result of land degradation (and may push people into
9 poverty), it can have further effects on issues such as migration (Lee 2009), as people adapt by
10 moving (see Section 4.7.3); and may result in conflict (see Section 4.7.3), as different groups within
11 society compete for scarce resources, sometimes through non-peaceful actions. Both migration and
12 conflict can lead to land use changes elsewhere that further fuel climate change through increased
13 emissions.

14 Similar challenges as for understanding land degradation-poverty linkages are experienced in
15 unravelling the relationship between climate change and poverty. A particular issue in examining
16 climate change-poverty links relates to the common use of aggregate economic statistics like GDP, as
17 the assets and income of the poor constitute such as minor proportion of national wealth (Hallegatte et
18 al. 2018). Aggregate quantitative measures also fail to capture the distributions of costs and benefits
19 from climate change. Furthermore, people fall into and out of poverty, with climate change being one
20 of many factors affecting these dynamics, through its impacts on livelihoods. Much of the literature
21 on climate change and poverty tends to look backward rather than forward (Skoufias et al. 2011),
22 providing a snap-shot of current or past relationships, (for example, (Dell et al. 2009) who examine
23 the relationship between temperature and income (GDP) using cross-sectional data from countries in
24 the Americas). Yet, simulations of future climate change impacts on income or poverty are largely
25 lacking.

26 Noting the *limited evidence* that exists that explicitly focuses on the relationship between land
27 degradation, climate change and poverty, Barbier and Hochard (2018b) suggest that those people
28 living in less favoured agricultural areas face a poverty-environment trap that can result in increased
29 land degradation under climate change conditions. The emergent relationships between land
30 degradation, climate change and poverty are shown in Figure 4.6 (see also Figure 6.1).

31



1
2 **Figure 4.6 Schematic representation of links between climate change, land management and socio-**
3 **economic conditions.**

4 The poor have access to few productive assets, so land, and the natural resource base more widely,
5 plays a key role in supporting the livelihoods of the poor. It is however, hard to make generalisations
6 about how important income derived from the natural resource base is for rural livelihoods in the
7 developing world (Angelsen et al. 2014) with studies focusing on forest resources having shown that
8 approximately one quarter of the total rural household income in developing countries stems from
9 forests, with forest-based income shares being tentatively higher for low-income households (Vedeld
10 et al. 2007; Angelsen et al. 2014). Different groups use land in different ways within their overall
11 livelihood portfolios and are therefore at different levels of exposure and sensitivity to climate shocks
12 and stresses. The literature nevertheless displays *high evidence* and *high agreement* that those
13 populations whose livelihoods are more sensitive to climate change and land degradation are often
14 more dependent on environmental assets, and these people are often the poorest members of society.
15 There is further *high evidence and high agreement* that both climate change and land degradation can
16 affect livelihoods and poverty through their threat multiplier effect. Research in Bellona, in the
17 Solomon Islands in the south Pacific (Reenberg et al. 2008) examined event-driven impacts on
18 livelihoods, taking into account weather events as one of many drivers of land degradation and links
19 to broader land-use and land cover changes that have taken place. Geographical locations
20 experiencing land degradation are often the same locations that are directly affected by poverty, and
21 are also affected by extreme events linked to climate change and variability.

22 Much of the assessment presented above has considered place-based analyses examining the
23 relationships between poverty, land degradation and climate change in the locations in which these
24 outcomes have occurred. Altieri and Nicholls (2017) note that due to the globalised nature of markets
25 and consumption systems, the impacts of changes in crop yields linked to climate-related land
26 degradation (manifest as lower yields) will be far reaching, beyond the sites and livelihoods
27 experiencing degradation. Despite these teleconnections, farmers living in poverty in developing
28 countries will be especially vulnerable due to their exposure, dependence on the environment for
29 income and limited options to engage in other ways to make a living (Rosenzweig and Hillel 1998). In
30 identifying ways in which these interlinkages can be addressed, (Scherr 2000) observes that key
31 actions that can jointly address poverty and environmental improvement often seek to increase access

1 to natural resources, enhance the productivity of the natural resource assets of the poor, and to engage
2 stakeholders in addressing public natural resource management issues. In this regard, it is increasingly
3 recognised that those suffering from and being vulnerable to land degradation and poverty need to
4 have a voice and play a role in the development of solutions, especially where the natural resources
5 and livelihood activities they depend on are further threatened by climate change.

6 **4.7.2 Impacts of climate related land degradation on food security**

7 How and where we grow food compared to where and when we need to consume it is at the crux of
8 issues surrounding land degradation, climate change and food security, especially because more than
9 75% of the global land surface (excluding Antarctica) faces rain-fed crop production constraints
10 (Fischer et al. 2009), see also Chapter 5. Taken separately, knowledge on land degradation processes
11 and human-induced climate change has attained a great level of maturity. However, their combined
12 effects on food security, notably food supply, remain underappreciated (Webb et al. 2017b), and
13 quantitative information is lacking. Just a few studies have shown how the interactive effects of the
14 aforementioned challenging, interrelated phenomena can impact crop productivity and hence food
15 security and quality (Karami et al. 2009; Allen et al. 2001; Högy and Fangmeier 2008) (*low evidence*).
16 Along with socio-economic drivers climate change accelerates land degradation due to its influence
17 on land-use systems (Millennium Assessment 2005; UNCCD 2017), potentially leading to a decline
18 in agri-food system productivity, particularly on the supply side. Increases in temperature and changes
19 in precipitation patterns are expected to have impacts on soil quality, including nutrient availability
20 and assimilation (St.Clair and Lynch 2010). Those climate-related changes are expected to have net
21 negative impacts on agricultural productivity, particularly in tropical regions, though the magnitude of
22 impacts depends on the models used. Anticipated supply side issues linked to land and climate relate
23 to biocapacity factors (including e.g. whether there is enough water to support agriculture); production
24 factors (e.g. chemical pollution of soil and water resources or lack of soil nutrients) and distribution
25 issues (e.g. decreased availability of and/or accessibility to the necessary diversity of quality food
26 where and when it is needed) (Stringer et al. 2011). Climate sensitive transport infrastructure is also
27 problematic for food security (Islam et al. 2017), and can lead to increased food waste, while poor
28 siting of roads and transport links can lead to soil erosion and forest loss (Xiao et al. 2017), further
29 feeding back into climate change.

30 Over the past decades, crop models have been useful tools for assessing and understanding climate
31 change impacts on crop productivity and food security (White et al. 2011; Rosenzweig et al. 2014).
32 Yet, the interactive effects of soil parameters and climate change on crop yields and food security
33 remain limited, with *low evidence* of how they play out in different economic and climate settings
34 (e.g. Sundström et al. 2014). Similarly, there have been few meta-analyses focusing on the adaptive
35 capacity of land-use practices such as conservation agriculture in light of climate stress (see e.g.
36 Steward et al. 2018), as well as *low evidence* quantifying the role of wild foods and forests (and by
37 extension forest degradation) in both the global food basket and in supporting household scale food
38 security (Bharucha and Pretty 2010; Hickey et al. 2016)

39 To be sustainable, any initiative aiming at addressing food security – encompassing supply, diversity
40 and quality - must take into consideration the interactive effects between climate and land degradation
41 in a context of other socio-economic stressors. Such socio-economic factors are especially important
42 if we look at demand side issues too, which include lack of purchasing power, large scale speculation
43 on global food markets leading to exponential price rises (Tadesse et al. 2014), competition in
44 appropriation of supplies and changes to per capita food consumption (Stringer et al. 2011; see also
45 Chapter 5). Lack of food security, combined with lack of livelihood options, is often an important
46 manifestation of vulnerability, and can act as a key trigger for people to migrate. In this way,
47 migration becomes an adaptation strategy.

1 **4.7.3 Impacts of climate-related land degradation on migration and conflict**

2 Land degradation may trigger competition for scarce natural resources potentially leading to
3 migration and/or conflict, though even with *medium evidence* there is *low agreement* in the literature.
4 Linkages between land degradation and migration occur within a larger context of multi-scale
5 interaction of environmental and non-environmental drivers and processes, including resettlement
6 projects, searches for education and/or income, land shortages, political turmoil, and family-related
7 reasons (McLeman 2017; Hermans and Ide 2019). The complex contribution of climate to migration
8 and conflict hampers retrieving any level of confidence on climate-migration and climate-conflict
9 linkages, therefore constituting a major knowledge gap (Cramer et al. 2014; Hoegh-Guldberg et al.
10 2018).

11 There is *low evidence* on the causal linkages between climate change, land degradation processes
12 (other than desertification) and migration. Existing studies on land degradation and migration –
13 particularly in drylands – largely focus on the effect of rainfall variability and drought and shows how
14 migration serves as adaptation strategy (Piguet et al. 2018; McLeman 2017; chapter 3). For example,
15 in the Ethiopian highlands severe topsoil erosion and forest degradation is a major environmental
16 stressor which is amplified by re-occurring droughts, with migration being an important household
17 adaptation strategy (Morrissey 2013). In the humid tropics, land degradation, mainly as a consequence
18 of deforestation, has been a reported reason for people leaving their homes during the Amazonian
19 colonisation (Hecht 1983) but was also observed more recently, for example in Guatemala, where soil
20 degradation was one of the most frequently cited migration pushes (López-Carr 2012) and Kenya,
21 where households respond to low soil quality by sending temporary migrants for additional income
22 generation (Gray 2011). In contrast, in the Andean highlands and the Pacific coastal plain, migration
23 increased with land quality, probably because revenues from additional agricultural production was
24 invested in costly forms of migration (Gray and Bilsborrow 2013). These mixed results illustrate the
25 complex, non-linear relationship of land degradation-migration linkages and suggest explaining land
26 degradation-migration linkages requires considering a broad socio-ecological embedding (McLeman
27 2017).

28 In addition to people moving away from an area due to “lost” livelihood activities, climate related
29 land degradation can also reduce the availability of livelihood safety nets – environmental assets that
30 people use during times of shocks or stress. For example, Barbier (2000) notes that wetlands in north-
31 east Nigeria around Hadejia–Jama’are floodplain provide dry season pastures for seminomadic
32 herders, agricultural surpluses for Kano and Borno states, groundwater recharge of the Chad
33 formation aquifer and ‘insurance’ resources in times of drought. The floodplain also supports many
34 migratory bird species. As climate change and land degradation combine, delivery of these multiple
35 services can be undermined, particularly as droughts become more widespread, reducing the utility of
36 this wetland environment as a safety net for people and wildlife alike.

37 Early studies conducted in Africa hint at a significant causal link between land degradation and
38 violent conflict (Homer-Dixon et al. 1993). For example, Percival and Homer-Dixon (1995) identified
39 land degradation as one of the drivers of the crisis in Rwanda in the early 1990s which allowed radical
40 forces to stoke ethnic rivalries. With respect to the Darfur conflict, some scholars and UNEP
41 concluded that land degradation, together with other environmental stressors, constitute a major
42 security threat for the Sudanese people (Byers and Dragojlovic 2004; Sachs 2007; UNEP 2007).
43 Recent studies show *low agreement*, suggesting that climate change can increase the likelihood of
44 civil violence if certain economic, political and social factors, including low development and weak
45 governance mechanisms, are present (Scheffran et al. 2012; Benjaminsen et al. 2012). In contrast,
46 Raleigh (Raleigh and Urdal 2007) found in a global study that land degradation is a weak predictor for
47 armed conflict. As such, studies addressing possible linkages between climate change – a key driver
48 of land degradation – and the risks of conflict have yielded contradictory results and it remains largely

1 unclear whether land degradation resulting from climate change leads to conflict or cooperation
2 (Salehyan 2008; Solomon et al. 2018).

3 Land degradation-conflict linkages can be bi-directional. Research suggests that households
4 experiencing natural resource degradation often engage in migration for securing livelihoods
5 (Kreamer 2012), which potentially triggers land degradation at the destination leading to conflict there
6 (Kassa et al. 2017). While this indeed holds true for some cases it may not for others given the
7 complexity of processes, contexts and drivers. Where conflict and violence do ensue, it is often as a
8 result of a lack of appreciation for the cultural practices of others.

9 **4.8 Addressing land degradation in the context of climate change**

10 Land degradation in the form of soil carbon loss is estimated to have been ongoing for at least 12,000
11 years, but increased exponentially in the last 200 years (Sanderman et al. 2017). Before the advent of
12 modern sources of nutrients, it was imperative for farmers to maintain and improve soil fertility
13 through the prevention of runoff and erosion, and management of nutrients through vegetation
14 residues and manure. Many ancient farming systems were sustainable for hundreds and even
15 thousands of years, such as raised field agriculture in Mexico (Crews and Gliessman 1991), tropical
16 forest gardens in SE Asia and Central America (Ross 2011; Torquebiau 1992; Turner and Sabloff
17 2012), terraced agriculture in East Africa, Central America, Southeast Asia and the Mediterranean
18 basin (Turner and Sabloff 2012; Preti and Romano 2014; Widgren and Sutton 2004; Håkansson and
19 Widgren 2007; Davies and Moore 2016; Davies 2015), and integrated rice-fish cultivation in East
20 Asia (Frei and Becker 2005).

21 Such long-term sustainable farming systems evolved in very different times and geographical
22 contexts, but they share many common features, such as: the combination of species and structural
23 diversity in time, and space (horizontally and vertically) in order to optimise the use of available land;
24 recycling of nutrients through biodiversity of plants, animals, and microbes; harnessing the full range
25 of site-specific micro-environments (e.g. wet and dry soils); biological interdependencies which helps
26 suppression of pests; reliance on mainly local resources; reliance on local varieties of crops and
27 sometimes incorporation of wild plants and animals; the systems are often labour and knowledge
28 intensive (Rudel et al. 2016; Beets 1990; Netting 1993; Altieri and Koohafkan 2008). Such farming
29 systems have stood the test of time and can provide important knowledge for adapting farming
30 systems to climate change (Koohafkan and Altieri 2011).

31 In modern agriculture the importance of maintaining the biological productivity and ecological
32 integrity of farm land has not been a necessity in the same way as in pre-modern agriculture because
33 nutrients and water have been supplied externally. The extreme land degradation in the US Midwest
34 during the Dust Bowl period in the 1930s became an important wake-up call for agriculture and
35 agricultural research and development, from which we can still learn much in order to adapt to
36 ongoing and future climate change (McLeman et al. 2014; Baveye et al. 2011; McLeman and Smit
37 2006).

38 Sustainable Land Management (SLM) is a unifying framework for addressing land degradation and
39 can be defined as the stewardship and use of land resources, including soils, water, animals and
40 plants, to meet changing human needs, while simultaneously ensuring the long-term productive
41 potential of these resources and the maintenance of their environmental functions. 'It is a
42 comprehensive approach comprising technologies combined with social, economic and political
43 enabling conditions (Nkonya et al. 2011). It is important to stress that farming systems are informed
44 by both scientific and local/traditional knowledge. The power of SLM in small-scale diverse farming
45 was demonstrated effectively in Nicaragua after the severe cyclone Mitch in 1998 (Holt-Giménez
46 2002). Pairwise analysis of 880 fields with and without implementation of SLM practices showed that

1 the SLM fields systematically fared better than the fields without SLM in terms of more topsoil
 2 remaining, higher field moisture, more vegetation, less erosion and lower economic losses after the
 3 cyclone. Furthermore the difference between fields with and without SLM increased with increasing
 4 levels of storm intensity, increasing slope gradient, and increasing age of SLM (Holt-Giménez 2002).

5 When addressing land degradation through SLM and other approaches it is important to consider
 6 feedbacks that impact climate change. Table 4.2 shows some of the most important land degradation
 7 issues, their potential solutions, and their impacts on climate change. This table provides a link
 8 between the comprehensive lists of land degradation processes (Table 4.1) and land management
 9 solutions (Table 4.2).

Table 4.2 (Cross-chapter Ch 3 and Ch 4) Interaction of human and climate drivers can exacerbate desertification and land degradation

Climate change exacerbates the rate and magnitude of several ongoing land degradation and desertification processes. Human drivers of land degradation and desertification include expanding agriculture, agricultural practices and forest management. In turn land degradation and desertification are also drivers of climate change through the emission of greenhouse gases, reduced rates of carbon uptake and reduced capacity of ecosystems to act as carbon sinks into the future.

| Human driver | Climate driver |
|--|---|
| Grazing pressure  | Warming trend  |
| Agriculture practices  | Extreme temperatures  |
| Expansion of agriculture  | Drying trend  |
| Forest clearing  | Extreme rainfall  |
| Wood fuel  | Shifting rains  |
| | Intensifying cyclones  |
| | Sea level rise  |

| Issue/syndrome | Impact on climate change | Human driver | Climate driver | Land management options | References |
|---|---|---|---|--|--|
| Erosion of agricultural soils | Emission: CO ₂ , N ₂ O |  |  | Increase soil organic matter, no till, perennial crops, erosion control, agro forestry, dietary change | {3.2.4, 3.5.1, 3.6.2, 3.8.1, 4.9.1, 4.9.5, 4.10.2, 4.10.5} |
| Deforestation | Emission of CO ₂ |  | | Forest protection, sustainable forest management and dietary change | {4.2.5, 4.6, 4.9.3, 4.9.4, 4.10.3} |
| Forest degradation | Emission of CO ₂ Reduced carbon sink |  | | Forest protection, sustainable forest management | {4.2.5, 4.6, 4.9.3, 4.9.4, 4.10.3} |
| Overgrazing | Emission: CO ₂ , CH ₄ Increasing albedo |  |  | Controlled grazing, rangeland management | {3.2.4.2, 3.5.1, 3.7.1, 3.8.1, 4.9.1.4} |
| Firewood and charcoal production | Emission: CO ₂ , CH ₄ Increasing albedo |  | | Clean cooking (health co-benefits, particularly for women and children) | {3.7.3, 4.6.4, 4.9.3, 4.9.4} |
| Increasing fire frequency and intensity | Emission: CO ₂ , CH ₄ , N ₂ O Emission: aerosols, increasing albedo | |  | Fuel management, fire management | {3.2.4, 3.7.1, 4.2.5, 4.9.3, Cross chapter box 3} |
| Degradation of tropical peat soils | Emission: CO ₂ , CH ₄ |  |  | Peatland restoration, erosion control, regulating the use of peat soils | {4.10.4} |
| Thawing of perma-frost | Emission: CO ₂ , CH ₄ | |  | relocation of settlement and infrastructure | {4.9.5.1} |
| Coastal erosion | Emission: CO ₂ , CH ₄ | |  | Wetland and coastal restoration, mangrove conservation, long term land use planning | {4.10.6, 4.10.7, 4.10.8} |
| Sand and dust storms, wind erosion | Emission: aerosols | |  | Vegetation management, afforestation, windbreaks | {3.4.1, 3.5.1, 3.7.1, 3.8.1, 3.8.2} |
| Bush encroachment | Capturing: CO ₂ , Decreasing albedo |  |  | Grazing land management, fire management | {3.7.1.3, 3.8.3.2} |

| |
|--|
| [3.1.4.3.4.1] [3.5.2.3.7.1.4.8.1.4] [8.5, 4.9.2.4.9.5] |
| [4.1.5.4.5.4.8.3.4.8] [4.4.9.3] |
| [4.1.5.4.5.4.8.3.4.8] [4.4.9.3] |
| [3.1.4.2.3.4.1.3.6.1] [3.7.1.4.8.1.4] |
| [3.6.3.4.5.4.4.8.3.4] [8.4] |
| [3.1.4.3.6.1.4.1.5.4] [8.3, Cross chapter box 3] |
| [4.9.4] |
| [4.8.5.1] |
| [4.9.6.4.9.7.4.9.8] |
| [3.3.1.3.4.1.3.6.1.3] [7.1.3.7.2] |
| [3.6.1.3.3.7.3.2] |

11
12

13 **4.8.1 Actions on the ground to address land degradation**

14 Concrete actions on the ground to address land degradation are primarily focused on soil and water
 15 conservation. In the context of adaptation to climate change, actions relevant for addressing land
 16 degradation are sometimes framed as ecosystem based adaptation (EBA) (Scarano 2017) or Nature
 17 Based Solutions (NBS) (Nesshöver et al. 2017), and in an agricultural context, agroecology (see

1 glossary) provides an important frame. The site-specific biophysical and social conditions, including
2 local and indigenous knowledge, are important for successful implementation of concrete actions.

3 Responses to land degradation generally take the form of agronomic measures (methods related to
4 managing the vegetation cover), soil management (methods related to tillage, nutrient supply), and
5 mechanical methods (methods resulting in durable changes to the landscape) (Morgan 2005a).
6 Measures may be combined to reinforce benefits to land quality, as well as improving carbon
7 sequestration that supports climate change mitigation. Some measures offer adaptation options and
8 other co-benefits, such as agroforestry involving planting fruit trees that can support food security in
9 the face of climate change impacts (Reed and Stringer 2016a) or application of compost or biochar
10 that enhances soil water holding capacity, so increases resilience to drought.

11 There are important differences in terms of labour and capital requirements for different technologies,
12 and also implications for land tenure arrangements. Agronomic measures and soil management
13 require generally little extra capital input and comprise activities repeated annually, so have no
14 particular implication for land tenure arrangements. Mechanical methods require substantial upfront
15 investments in terms of capital and labour, resulting in long lasting structural change requiring more
16 secure land tenure arrangements (Mekuriaw et al. 2018). Agroforestry is a particularly important
17 strategy for SLM in the context of climate change because the large potential to sequester carbon in
18 plants and soil and enhance resilience of agricultural systems (Zomer et al. 2016).

19 Implementation of sustainable land management practices has been shown to increase the productivity
20 of land (Branca et al. 2013) and to provide good economic returns on investment in many different
21 settings around the world (Mirzabaev et al. 2015). Giger et al (2018) showed in a meta study of 363
22 projects over the period 1990 to 2012 that 73% of the projects were perceived to have a positive or at
23 least neutral cost/benefit ratio in the short term, and 97% were perceived to have a positive or very
24 positive cost/benefit ratio in the long term (*robust evidence, high agreement*). Despite the positive
25 effects, uptake is far from universal. Local factors, both biophysical conditions (e.g. soils, drainage,
26 and topography) and socio-economic conditions (e.g. land tenure, economic status, and land
27 fragmentation) play decisive roles in the interest in, capacity to undertake, and successful
28 implementation of sustainable land management practices (Teshome et al. 2016; Vogl et al. 2017;
29 Tesfaye et al. 2016; Cerdà et al. 2018; Adimassu et al. 2016). From a landscape perspective,
30 sustainable land management can generate benefits, including adaptation to and mitigation of climate
31 change, for entire watersheds, but challenges remain regarding coordinated and consistent
32 implementation (Kerr et al. 2016; Wang et al. 2016a). (*medium evidence, medium agreement*)

33 **4.8.1.1 Agronomic and soil management measures**

34 Rebuilding soil carbon is an important goal of SLM, particularly in the context of climate change
35 (Rumpel et al. 2018). The two most important reasons why agricultural soils have lost 20-60% of the
36 soil carbon they contained under natural ecosystem conditions are the frequent disturbance through
37 tillage and harvesting and the change from deep rooted perennial plants to shallow rooted annual
38 plants (Crews and Rumsey 2017). Practices that build soil carbon are those that increase organic
39 matter input to soil, or reduce decomposition of soil organic matter.

40 Agronomic practices can alter the carbon balance significantly, by increasing organic inputs from
41 litter and roots into the soil. Practices include retention of residues, use of locally-adapted varieties,
42 inter-cropping, crop rotations, and green manure crops that replace the bare field fallow during winter
43 and are eventually ploughed before sowing next main crop (Henry et al., 2018). Cover crops (green
44 manure crops and catch crops that are grown between the main cropping seasons) can increase soil
45 carbon stock by between 0.22 and 0.4 t C ha⁻¹yr⁻¹ (Poepflau and Don 2015; Kaye and Quemada 2017).

46 Reduced tillage (or no-tillage) is an important strategy for reducing soil erosion and nutrient loss by
47 wind and water (Van Pelt et al. 2017; Panagos et al. 2015; Borrelli et al. 2016). But the evidence that

1 no-till agriculture also sequesters carbon is not compelling (VandenBygaart 2016). Soil sampling of
2 only the upper 30 cm can give biased results suggesting that soils under no-till practices have higher
3 carbon content than soils under conventional tillage (Baker et al. 2007; Ogle et al. 2012; Fargione et
4 al. 2018; VandenBygaart 2016).

5 Changing from annual to perennial crops can increase soil carbon content (Culman et al. 2013; Sainju
6 et al. 2017). A perennial grain crop (intermediate wheatgrass) was on average over four years a net
7 carbon sink of about 13.5 t CO₂ ha⁻¹yr⁻¹ (de Oliveira et al. 2018). Sprunger et al. (2018) compared an
8 annual winter wheat crop with a perennial grain crop (intermediate wheatgrass) and found that the
9 perennial grain root biomass was 15 times larger than winter wheat, however, there was no significant
10 difference in soil carbon pools after the four-year experiment. Exactly how much, and over what time
11 period, carbon can be sequestered through changing from annual to perennial crops depends on the
12 degree of soil carbon depletion and other local biophysical factors (see also section 4.9.2).

13 Integrated soil fertility management is a sustainable approach to nutrient management that uses a
14 combination of chemical and organic amendments (manure, compost, biosolids, biochar), rhizobial
15 nitrogen fixation, and liming materials to address soil chemical constraints (Henry et al., 2018). In
16 pasture systems, management of grazing pressure, fertilisation, diverse species including legumes and
17 perennial grasses can reduce erosion and enhance soil carbon (Conant et al. 2017).

18 **4.8.1.2 Mechanical soil and water conservation**

19 In hilly and mountainous terrain terracing is an ancient but still practiced soil conservation method
20 worldwide (Preti and Romano 2014) in climatic zones from arid to humid tropics (Balbo 2017). By
21 reducing the slope gradient of hillsides, terraces provide flat surfaces and deep, loose soils that
22 increase infiltration, reduce erosion and thus sediment transport. They also decrease the hydrological
23 connectivity and thus reduce hillside runoff (Preti et al. 2018; Wei et al. 2016; Arnáez et al. 2015;
24 Chen et al. 2017). In terms of climate change, terraces are a form of adaptation which helps both in
25 cases where rainfall is increasing or intensifying (by reducing slope gradient and the hydrological
26 connectivity), and where rainfall is decreasing (by increasing infiltration and reducing runoff) (*robust
27 evidence, high agreement*). There are several challenges, however, to continued maintenance and
28 construction of new terraces, such as the high costs in terms of labour and/or capital (Arnáez et al.
29 2015) and disappearing local knowledge for maintaining and constructing new terraces (Chen et al.
30 2017). The propensity of farmers to invest in mechanical soil conservation methods varies with land
31 tenure, farmers with secure tenure arrangements are more willing to invest in durable practices such
32 as terraces (Lovo 2016; Sklenicka et al. 2015; Haregeweyn et al. 2015). Where the slope is less
33 severe, erosion can be controlled by contour banks, and the keyline approach (Duncan 2016; Stevens
34 et al. 2015) to soil and water conservation.

35 **4.8.1.3 Agroforestry**

36 Agroforestry is defined as a collective name for land-use systems in which woody perennials (trees,
37 shrubs, etc.) are grown in association with herbaceous plants (crops, pastures) and/or livestock in a
38 spatial arrangement, a rotation, or both, and in which there are both ecological and economic
39 interactions between the tree and non-tree components of the system (Young, 1995, p. 11). At least
40 since the 1980s agroforestry has been widely touted as an ideal land management practice in areas
41 vulnerable to climate variations and subject to soil erosion. Agroforestry holds the promise of
42 improving of soil and climatic conditions while generating income from wood energy, timber, and
43 non-timber products – sometimes presented as a synergy of adaptation and mitigation of climate
44 change (Mbow et al. 2014).

45 There is strong scientific consensus that a combination of forestry with agricultural crops and/or
46 livestock, agroforestry systems can provide additional ecosystem services when compared with
47 monoculture crop systems (Waldron et al. 2017; Sonwa et al. 2011a, 2014, 2017; Charles et al. 2013).
48 Agroforestry can enable sustainable intensification by allowing continuous production on the same

1 unit of land with higher productivity without the need to use shifting agriculture systems to maintain
2 crop yields (Nath et al. 2016). This is especially relevant where there is a regional requirement to find
3 a balance between the demand for increased agricultural production and the protection of adjacent
4 natural ecosystems such as primary and secondary forests (Mbow et al. 2014). For example, the use of
5 agroforestry for perennial crops such as coffee and cocoa are increasingly promoted as offering a
6 route to sustainable farming with important climate change adaptation and mitigation co-benefits
7 (Sonwa et al. 2001; Kroeger et al. 2017). Reported co-benefits of agroforestry in cocoa production
8 include increased carbon sequestration in soils and biomass, improved water and nutrient use
9 efficiency and the creation of a favourable micro-climate for crop production (Sonwa et al. 2017; Chia
10 et al. 2016). Importantly, the maintenance of soil fertility using agroforestry has the potential to
11 reduce the practice of shifting-agriculture (of cocoa) which results in deforestation (Gockowski and
12 Sonwa 2011). However, positive interactions within these systems can be ecosystem and/or species
13 specific, but co-benefits such as increased resilience to extreme climate events, or improved soil
14 fertility are not always observed (Blaser et al. 2017; Abdulai et al. 2018). These contrasting outcomes
15 indicate the importance of field scale research programs to inform agroforestry system design, species
16 selection and management practices (Sonwa et al. 2014) .

17 Despite the many proven benefits, adoption of agroforestry has been low and slow (Toth et al. 2017;
18 National Research Centre for Agroforestry et al. 1999; Pattanayak et al. 2003; Jerneck and Olsson
19 2014). There are several reasons for the slow uptake, but the perception of risks and the time lag
20 between adoption and realisation of benefits are often important (Pattanayak et al. 2003; Mercer 2004;
21 Jerneck and Olsson 2013).

22 An important question for agroforestry is whether it supports poverty alleviation, or if it favours
23 comparatively affluent households. Experiences from India suggest that the overall adoption is (s)low
24 and differential between rich and poor households. Brockington et al. (2016), studied agroforestry
25 adoption over many years in South India, they found that overall only 18% of the households adopted
26 agroforestry but among the relatively rich households who adopted agroforestry, 97% of them were
27 still practicing it after 6-8 years and some had expanded their operations. Similar results were
28 obtained in Western Kenya, that food secure households were much more willing to adopt
29 agroforestry than food insecure households (Jerneck and Olsson 2013, 2014). Other experiences from
30 sub-Saharan Africa illustrate the difficulties (such as local institutional support) of having a continued
31 engagement of communities in agroforestry (Noordin et al. 2001; Matata et al. 2013; Meijer et al.
32 2015).

33 ***4.8.1.4 Crop-livestock interaction as an approach to manage land degradation***

34 The integration of crop and livestock production into “mixed farming” for smallholders in developing
35 countries became an influential model, particularly for Africa, in the early 1990s (Pritchard et al.
36 1992; McIntire et al. 1992). Crop-livestock integration under this model was seen as founded on three
37 pillars; improved use of manure for crop fertility management; expanded use of animal traction
38 (draught animals); and promotion of cultivated fodder crops. For Asia, emphasis was placed on
39 draught power for land preparation, manure for soil fertility enhancement, and fodder production as
40 an entry point for cultivation of legumes (Devendra and Thomas 2002). Mixed farming was seen as an
41 evolutionary process to expand food production in the face of population increase, promote
42 improvements in income and welfare, and protect the environment. The process could be further
43 facilitated and steered by research, extension and policy (Pritchard et al. 1992; McIntire et al. 1992;
44 Devendra 2002) (Pritchard et al., 1992; McIntire et al. 1992; Devendra 1992).

45 Scoones and Wolmer (2002) place this model in historical context, including concern about
46 population pressure on resources and the view that mobile pastoralism was environmentally
47 damaging. The latter view had already been critiqued by developing understandings of pastoralism,
48 mobility and communal tenure of grazing lands (for example (Behnke 1994; Ellis 1994)). They set out

1 a much more differentiated picture of crop livestock interactions, which can take place either within a
2 single farm household, or between crop and livestock producers, in which case they will be mediated
3 by formal and informal institutions governing the allocation of land, labour and capital, with the
4 interactions evolving through multiple place-specific pathways (Ramisch et al. 2002; Scoones and
5 Wolmer 2002). Promoting a diversity of approaches to crop-livestock interactions does not imply that
6 the integrated model necessarily leads to land degradation, but increases the space for institutional
7 support to local innovation (Scoones and Wolmer 2002).

8 However, specific managerial and technological practices that link crop and livestock production will
9 remain an important part of the repertoire of on-farm adaptation and mitigation. Howden and
10 coauthors (Howden et al. 2007) note the importance of innovation within existing integrated systems
11 including use of adapted forage crops. Rivera-Ferre et al. (2016) list as adaptation strategies with
12 high potential for grazing systems, mixed crop-livestock systems or both: crop-livestock integration in
13 general; soil management including composting; enclosure and corralling of animal; improved storage
14 of feed. Most of these are seen as having significant co-benefits for mitigation, and improved
15 management of manure is seen as a mitigation measure with adaptation co-benefits.

16 **4.8.2 Local and indigenous knowledge for addressing land degradation**

17 In practice, responses are anchored both in scientific research, as well as local, indigenous and
18 traditional knowledge and know-how. For example, studies in the Philippines Camacho et al. (2016)
19 examine how traditional integrated watershed management by indigenous people sustain regulating
20 services vital to agricultural productivity, while delivering co-benefits in the form of biodiversity and
21 ecosystem resilience at a landscape scale. Although responses can be site specific and sustainable at a
22 local scale, the multi-scale interplay of drivers and pressures can nevertheless cause practices that
23 have been sustainable for centuries to become less so. Siahaya et al (2016) explore the traditional
24 knowledge that has informed rice cultivation in the uplands of East Borneo, grounded in sophisticated
25 shifting cultivation methods (gilir balik) which have been passed on for generations (more than 200
26 years) in order to maintain local food production. Gilir balik involves temporary cultivation of plots,
27 after which, abandonment takes place as the land user moves to another plot, leaving the natural
28 (forest) vegetation to return. This approach is considered sustainable if it has the support of other
29 subsistence strategies, adapts to and integrates with the local context, and if the carrying capacity of
30 the system is not surpassed (Siahaya et al. 2016). Often gilir balik cultivation involves intercropping
31 of rice with bananas, cassava and other food crops. Once the abandoned plot has been left to recover
32 such that soil fertility is restored, clearance takes place again and the plot is reused for cultivation.
33 Rice cultivation in this way plays an important role in forest management, with several different types
34 of succession forest being found in the study area of Siahaya et al (2016). Nevertheless, interplay of
35 these practices with other pressures (large-scale land acquisitions for oil palm plantation, logging and
36 mining), risk their future sustainability. Use of fire is critical in processes of land clearance, so there
37 are also trade-offs for climate change mitigation which have been sparsely assessed.

38 Interest appears to be growing in understanding how indigenous and local knowledge inform land
39 users' responses to degradation, as scientists engage farmers as experts in processes of knowledge co-
40 production and co-innovation (Oliver et al. 2012; Bitzer and Bijman 2015). This can help to
41 introduce, implement, adapt and promote the use of locally appropriate responses (Schwilch et al.
42 2011). Indeed, studies strongly agree on the importance of engaging local populations in both
43 sustainable land and forest management. Meta-analyses in tropical regions that examined both forests
44 in protected areas and community managed forests suggest that deforestation rates are lower, with less
45 variation in deforestation rates presenting in community managed forests compared to protected
46 forests (Porter-Bolland et al. 2012). This suggests that consideration of the social and economic needs
47 of local human populations is vital in preventing forest degradation (Ward et al. 2018). However,
48 while disciplines such as ethnopedology seek to record and understand how local people perceive,

1 classify and use soil, and draw on that information to inform its management (Barrera-Bassols and
2 Zinck 2003), links with climate change and its impacts (perceived and actual) are not generally
3 considered.

4 **4.8.3 Reducing deforestation and forest degradation and increasing afforestation**

5 Improved stewardship of forests through reduction or avoidance of deforestation and forest
6 degradation, and enhancement of forest carbon stocks can all contribute to land-based natural climate
7 solutions (Angelsen et al. 2018; Sonwa et al. 2011b; Griscom et al. 2017). While estimates of annual
8 emissions from tropical deforestation and forest degradation range widely from 0.5 to 3.5 Gt C yr⁻¹
9 (Baccini et al. 2017; Houghton et al. 2012; Mitchard 2018, see also Chapter 2), they all indicate the
10 large potential to reduce annual emissions from deforestation and forest degradation. Recent estimates
11 of forest extent for Africa in 1900 may result in downward adjustments of historic deforestation and
12 degradation emission estimates (Aleman et al. 2018). Emissions from forest degradation in non-
13 Annex I countries have declined marginally from 1.1 GtCO₂ yr⁻¹ in 2001-2010 to 1 GtCO₂ yr⁻¹ in
14 2011-2015, but the relative emissions from degradation compared to deforestation have increased
15 from a quarter to a third (Federici et al. 2015). Forest sector activities in developing countries were
16 estimated to represent a technical mitigation potential in 2030 of 9 Gt CO₂ (Miles et al. 2015). This
17 was partitioned into reduction of deforestation (3.5 Gt CO₂), reduction in degradation and forest
18 management (1.7 Gt CO₂) and afforestation and reforestation (3.8 GtCO₂). The economic mitigation
19 potential will be lower than the technical potential (Miles et al. 2015).

20 Natural regeneration of second-growth forests enhances carbon sinks in the global carbon budget
21 (Chazdon and Uriarte 2016). In Latin America, Chazdon et al. (2016) estimated that in 2008, second-
22 growth forests (1 to 60 years old) covered 2.4 M km² of land (28.1% of the total study area). Over 40
23 years, these lands can potentially accumulate 8.5 Gt C in aboveground biomass via low-cost natural
24 regeneration or assisted regeneration, corresponding to a total CO₂ sequestration of 31.1 Gt CO₂
25 (Chazdon et al. 2016b). While aboveground biomass carbon stocks are estimated to be declining in
26 the tropics, they are increasing globally due to increasing stocks in temperate and boreal forests (Liu
27 et al. 2015b), consistent with the observations of a global land sector carbon sink (Le Quéré et al.
28 2013; Keenan et al. 2017; Pan et al. 2011).

29 Moving from technical mitigation potentials (Miles et al. 2015) to real reduction of emissions from
30 deforestation and forest degradation required transformational changes (Korhonen-Kurki et al. 2018).
31 This transformation can be facilitated by two enabling conditions: the presence of already initiated
32 policy change; or the scarcity of forest resources combined with an absence of any effective forestry
33 framework and policies. These authors and others (Angelsen et al. 2018) found that the presence of
34 powerful transformational coalitions of domestic pro-REDD+ political actors combined with strong
35 ownership and leadership, regulations and law enforcement, and performance-based funding, can
36 provide a strong incentive for achieving REDD+ goals.

37 Implementing schemes such as REDD+ and various projects related to the voluntary carbon market is
38 often regarded as a no-regrets investment (Seymour and Angelsen 2012) but the social and ecological
39 implications (including those identified in the Cancun Safeguards) must be carefully considered for
40 REDD+ projects to be socially and ecologically sustainable (Jagger et al. 2015). In 2018, 34 countries
41 have submitted a REDD+ forest reference level and/or forest reference emission level to the
42 UNFCCC. Of these REDD+ reference levels, 95% included the activity "reducing deforestation"
43 while 34% included the activity "reducing forest degradation" (FAO 2018). Five countries submitted
44 REDD+ results in the technical annex to their Biannual Update Report (BUR) totalling an emission
45 reduction of 6.3 Gt CO₂ between 2006 and 2015 (FAO 2018).

46 Afforestation is another mitigation activity that increases carbon sequestration (see also Cross-Chapter
47 Box 2: Implications of large-scale reforestation and afforestation, Chapter 1). Yet, it requires careful

1 consideration about where to plant trees to achieve potential climatic benefits given an altering of
2 local albedo and turbulent energy fluxes and increasing night-time land surface temperatures (Peng et
3 al., 2014). A recent hydro-climatic modelling effort has shown that forest cover can account for about
4 40% of the observed decrease in annual runoff (Buendia et al. 2016). A meta-analysis of afforestation
5 in Northern Europe (Bárcena and co-authors 2014) concluded that significant soil organic carbon
6 sequestration in Northern Europe occurs after afforestation of croplands but not grasslands. Additional
7 sequestration occurs in forest floors and biomass carbon stocks. Successful programmes of large scale
8 afforestation activities in South Korea and China are discussed in-depth a special case study (Section
9 4.9.3).

10 The potential outcome of efforts to reduce emissions from deforestation and degradation in Indonesia
11 through a 2011 moratorium on concessions to convert primary forests to either timber or palm oil uses
12 was evaluated against rates of emissions over the period 2000 to 2010. The study concluded that less
13 than 7% of emissions would have been avoided had the moratorium been implemented in 2000
14 because it only curtailed emissions due to a subset of drivers of deforestation and degradation (Busch
15 et al. 2015).

16 In terms of ecological integrity of tropical forests, the policy focus on carbon storage and tree cover
17 can be problematic if it leaves out other aspects of forests ecosystems, such as biodiversity – and
18 particularly fauna (Panfil and Harvey 2016; Peres et al. 2016; Hinsley et al. 2015). Other concerns of
19 forest based projects under the voluntary carbon market are potential negative socio-economic side
20 effects (Edstedt and Carton 2018a; Carton and Andersson 2017; Osborne 2011; Scheidel and Work
21 2018; Richards and Lyons 2016; Borrás and Franco 2018; Paladino and Fiske 2017) and leakage
22 (particularly at the subnational scale), i.e. when interventions to reduce deforestation or degradation at
23 one site displace pressures and increase emissions elsewhere (Atmadja and Verchot 2012; Phelps et
24 al. 2010; Lund et al. 2017; Balooni and Lund 2014).

25 Maintaining and increasing forest area, in particular of native forests rather than monoculture and
26 short-rotation plantations, contributes to the maintenance of global forest carbon stocks (Lewis et al.
27 2019) (*robust evidence, high agreement*).

28 **4.8.4 Sustainable forest management and CO₂ removal technologies**

29 While reducing deforestation and forest degradation may help directly meet mitigation goals,
30 sustainable forest management aimed at providing timber, fiber, biomass and non-timber resources
31 can provide long-term livelihood for communities, can reduce the risk of forest conversion to non-
32 forest uses (settlement, crops, etc.), and can maintain land productivity, thus reducing the risks of land
33 degradation (Putz et al. 2012; Gideon Neba et al. 2014; Sufo Kankeu et al. 2016; Nitcheu Tchiadje et
34 al. 2016; Rossi et al. 2017).

35 Developing sustainable forest management strategies aimed at contributing towards negative
36 emissions throughout this century requires an understanding of forest management impacts on
37 ecosystem carbon stocks (including soils), carbon sinks, carbon fluxes in harvested wood, carbon
38 storage in harvested wood products including landfills and the emission reductions achieved through
39 the use of wood products and bioenergy (Nabuurs et al. 2007; Lemprière et al. 2013; Kurz et al. 2016;
40 Law et al. 2018; Nabuurs et al. 2017). Transitions from natural to managed forest landscapes can
41 involve a reduction in forest carbon stocks, the magnitude of which depends on the initial landscape
42 conditions, the harvest rotation length relative to the frequency and intensity of natural disturbances
43 and on the age-dependence of managed and natural disturbances (Harmon et al. 1990; Kurz et al.
44 1998a). Initial landscape conditions, in particular the age-class distribution and therefore C stocks of
45 the landscape strongly affect the mitigation potential of forest management options (Ter-Mikaelian et
46 al. 2013; Kilpeläinen et al. 2017). Landscapes with predominantly mature forests may experience
47 larger reductions in carbon stocks during the transition to managed landscapes (Harmon et al. 1990;

1 Kurz et al. 1998b; Lewis et al. 2019) while in landscapes with predominantly young or recently
2 disturbed forests sustainable forest management can enhance carbon stocks (Henttonen et al. 2017).

3 Forest growth rates, net primary productivity, and net ecosystem productivity are age-dependent with
4 maximum rates of carbon removal from the atmosphere occurring in young to medium aged forests
5 and declining thereafter (Tang et al. 2014). In boreal forest ecosystem, estimation of carbon stocks
6 and carbon fluxes indicate that old growth stands are typically small carbon sinks or carbon sources
7 (Gao et al. 2018; Taylor et al. 2014; Hadden and Grelle 2016). In tropical forests, carbon uptake rates
8 in the first 20 years of forest recovery were 11 times higher than uptake rates in old-growth forests
9 (Poorter et al. 2016). Age-dependent increases in forest carbon stocks and declines in forest carbon
10 sinks mean that landscapes with older forests have accumulated more carbon but their sink strength is
11 diminishing, while landscapes with younger forests contain less carbon but they are removing CO₂
12 from the atmosphere at a much higher rate (Volkova et al. 2017; Poorter et al. 2016). The rates of
13 carbon removal are not just age-related but also controlled by many biophysical factors and human
14 activities (Bernal et al. 2018) and in ecosystems with uneven-aged, multispecies forests the
15 relationships between carbon stocks and sinks are more difficult and expensive to quantify.

16 Whether or not forest harvest and use of biomass is contributing to net reductions of atmospheric
17 carbon depends on carbon losses during and following harvest, rates of forest regrowth, and the use of
18 the harvested wood and the carbon retention in long-lived or short-lived products as well as the
19 emission reductions achieved through the substitution of emissions-intensive products with wood
20 products (Lemprière et al. 2013; Lundmark et al. 2014; Xu et al. 2018b; Olguin et al. 2018; Dugan
21 et al. 2018; Chen et al. 2018b; Pingoud et al. 2018; Seidl et al. 2007). Studies that ignore changes in
22 forest carbon stocks (such as some life cycle analyses that assume no impacts of harvest on forest
23 carbon stocks), ignore changes in wood product pools (Mackey et al. 2013) or assume long-term
24 steady state (Pingoud et al. 2018), or ignore changes in emissions from substitution benefits (Mackey
25 et al. 2013; Lewis et al. 2019) will arrive at diverging conclusions about the benefits of sustainable
26 forest management. Moreover, assessments of climate benefits of any mitigation action must also
27 consider the time dynamics of atmospheric impacts as some actions will have immediate benefits (e.g.
28 avoided deforestation) while others may not achieve net atmospheric benefits for decades or centuries.
29 For example, the climate benefits of woody biomass use for bioenergy depend on several factors such
30 as the source and alternate fate of the biomass, the energy type it substitutes and the rates of regrowth
31 of the harvested forest (Laganière et al. 2017; Ter-Mikaelian et al. 2014; Smyth et al. 2017).
32 Conversion of primary forests in regions of very low stand replacing disturbances to short-rotation
33 plantations where the harvested wood is used for short-lived products with low displacement factors
34 will increase emissions. In general, greater mitigation benefits are achieved if harvested wood
35 products are used for products with long carbon retention time and high displacement factors.

36 With increasing forest age, carbon sinks in forests will diminish until harvest or natural disturbances
37 such as wildfire remove biomass carbon or release it to the atmosphere (Seidl et al. 2017). While
38 individual trees can accumulate carbon for centuries (Köhl et al. 2017), stand level carbon
39 accumulation rates depend on both tree growth and tree mortality rates (Hember et al. 2016; Lewis et
40 al. 2004). Sustainable forest management, including harvest and forest regeneration, can help
41 maintain active carbon sinks by maintaining a forest age-class distribution that includes a share of
42 young, actively growing stands (Volkova et al. 2018; Nabuurs et al. 2017). The use of the harvested
43 carbon in either long-lived wood products (e.g. for construction), short-lived wood products (e.g.,
44 pulp and paper), or biofuels affects the net carbon balance of the forest sector (Lemprière et al. 2013;
45 Matthews et al. 2018). The use of these wood products can further contribute to GHG emission
46 reduction goals by avoiding the emissions from the products with higher embodied emissions that
47 have been displaced (Nabuurs et al. 2007; Lemprière et al. 2013). In 2007 the IPCC concluded that
48 “[i]n the long term, a sustainable forest management strategy aimed at maintaining or increasing

1 forest carbon stocks, while producing an annual sustained yield of timber, fibre or energy from the
2 forest, will generate the largest sustained mitigation benefit” (Nabuurs et al. 2007).–The apparent
3 trade-offs between maximising forest C stocks and maximising ecosystem C sinks are at the origin of
4 ongoing debates about optimum management strategies to achieve negative emissions (Keith et al.
5 2014; Kurz et al. 2016; Lundmark et al. 2014). Sustainable forest management, including the
6 intensification of carbon-focussed management strategies, can make long-term contributions towards
7 negative emissions if the sustainability of management is assured through appropriate governance,
8 monitoring and enforcement. As specified in the definition of sustainable forest management, other
9 criteria such as biodiversity must also be considered when assessing mitigation outcomes (Lecina-
10 Diaz et al. 2018). Moreover, the impacts of changes in management on albedo and other non-GHG
11 factors also need to be considered (Luyssaert et al. 2018) (See also Chapter 2). The contribution of
12 sustainable forest management for negative emissions is strongly affected by the use of the wood
13 products derived from forest harvest and the time horizon over which the carbon balance is assessed.
14 Sustainable forest management needs to anticipate the impacts of climate change on future tree
15 growth, mortality and disturbances when designing climate change mitigation and adaptation
16 strategies (Valade et al. 2017; Seidl et al. 2017).

17 **4.8.5 Policy responses to land degradation**

18 The 1992 United Nations Conference on Environment and Development (UNCED), also known as
19 the Rio de Janeiro Earth Summit, recognised land degradation as a major challenge to sustainable
20 development, and led to the establishment of the United Nations Convention to Combat
21 Desertification (UNCCD), which addressed specifically land degradation in the drylands. The
22 UNCCD emphasizes sustainable land use to link poverty reduction on one hand and environmental
23 protection on the other. The two other “Rio Conventions” emerging from the UNCED, the United
24 Nations Framework Convention on Climate Change (UNFCCC) and the Convention on Biological
25 Diversity (CBD), focus on climate change and biodiversity, respectively. The land has been
26 recognized as an aspect of common interest to the three conventions, and sustainable land
27 management (SLM) is proposed as a unifying theme for current global efforts on combating land
28 degradation, climate change and loss of biodiversity, as well as facilitating land-based adaptation to
29 climate change and sustainable development.

30 The Global Environmental Facility (GEF) funds developing countries to undertake activities that meet
31 the goals of the conventions and deliver global environmental benefits. Since 2002, the GEF has
32 invested in projects that support sustainable land management through its Land Degradation Focal
33 Area Strategy, to address land degradation within and beyond the drylands.

34 Under the UNFCCC, parties have devised National Adaptation Plans (NAPs) that identify medium-
35 and long-term adaptation needs. Parties have also developed their climate change mitigation plans,
36 presented as Nationally Determined Contributions (NDCs). These programs have the potential of
37 assisting the promotion of SLM. It is realised that the root causes of land degradation and successful
38 adaptation will not be realised until holistic solutions to land management are explored. SLM can help
39 address root causes of low productivity, land degradation, loss of income generating capacity as well
40 as contribute to the amelioration of the adverse effects of climate change.

41 The “4 per 1000” (4p1000) initiative (Soussana et al. 2019) launched by France during the UNFCCC
42 COP21 in 2015 aims at capturing CO₂ from the atmosphere through changes to agricultural and
43 forestry practices at a rate that would increase the carbon content of soils by 0.4% per year (Rumpel et
44 al. 2018). If global soil carbon content increases at this rate in the top 30-40 cm, the annual increase in
45 atmospheric CO₂ would be stopped (Dignac et al. 2017). This is an illustration of how extremely
46 important soils are for addressing climate change. The initiative is based on eight steps: stop carbon
47 loss (priority #1 is peat soils); promote carbon uptake; monitor, report, and verify impacts; deploy

1 technology for tracking soil carbon; test strategies for implementation and upscaling; involve
2 communities; coordinate policies; provide support (Rumpel et al. 2018). Questions remain however,
3 to what extent the 4p1000 is achievable as a universal goal (van Groenigen et al. 2017; Poulton et al.
4 2018; Schlesinger and Amundson 2018).

5 Land degradation neutrality (LDN) was introduced by the UNCCD at Rio +20, and adopted at
6 UNCCD COP12 (UNCCD 2016a). LDN is defined as "a state whereby the amount and quality of land
7 resources necessary to support ecosystem functions and services and enhance food security remain
8 stable or increase within specified temporal and spatial scales and ecosystems". Pursuit of LDN
9 requires effort to avoid further net loss of the land-based natural capital relative to a reference state, or
10 baseline. LDN encourages a dual-pronged effort involving sustainable land management to reduce the
11 risk of land degradation, combined with efforts in land restoration and rehabilitation, to maintain or
12 enhance land-based natural capital, and its associated ecosystem services (Orr et al., 2017; Cowie et
13 al. 2018;). Planning for LDN involves projecting the expected cumulative impacts of land use and
14 land management decisions, then counterbalancing anticipated losses with measures to achieve
15 equivalent gains, within individual land types (where land type is defined by land potential). Under
16 LDN framework developed by UNCCD, three primary indicators are used to assess whether LDN is
17 achieved by 2030: land cover change, net primary productivity and soil organic carbon (Cowie et al.
18 2018; Sims et al., 2019. Achieving LDN therefore requires integrated landscape management that
19 seeks to optimize land use to meet multiple objectives (ecosystem health, food security, human well-
20 being) (Cohen-Shacham, E., Walters, G., Janzen, C. and Maginnis 2016). The response hierarchy of
21 Avoid > Reduce > Reverse land degradation articulates the priorities in planning LDN interventions.
22 LDN provides the impetus for widespread adoption of SLM and efforts to restore or rehabilitate land.
23 Through its focus LDN ultimately provides tremendous potential for mitigation of and adaptation to
24 climate change by halting and reversing land degradation and transforming land from a carbon source
25 to a sink. There are strong synergies between the concept of LDN and the Nationally Determined
26 Contributions (NDCs) of many countries with linkages to national climate plans. LDN is also closely
27 related to many Sustainable Development Goals (SDG) in the areas of poverty, food security,
28 environmental protection and sustainable use of natural resources (UNCCD 2016b). The GEF is
29 supporting countries to set LDN targets and implement their LDN plans through its land degradation
30 focal area, which encourages application of integrated landscape approach to managing land
31 degradation (GEF 2018).

32 The 2030 agenda for sustainable development, adopted by the United Nations in 2015, comprises 17
33 Sustainable Development Goals (SDGs). Goal 15 is of direct relevance to land degradation with the
34 objective to protect, restore and promote sustainable use of terrestrial ecosystems, sustainably manage
35 forests, combat desertification and halt and reverse land degradation and halt biodiversity loss. Target
36 15.3 specifically addresses land degradation neutrality. Other goals that are relevant for land
37 degradation include goal 2 (Zero hunger), goal 3 (Good health and well-being), goal 7 (Affordable
38 and clean energy), goal 11 (Sustainable cities and communities), and goal 12 (Responsible production
39 and consumption). Sustainable management of land resources underpins the SDGs related to hunger,
40 climate change and environment. Further goals of a cross-cutting nature include 1 (No poverty), 6
41 (Clean water and sanitation) and 13 (Climate action). It remains to be seen how these interconnections
42 are dealt with in practice.

43 With a focus on biodiversity, IPBES published a comprehensive assessment of land degradation in
44 2018 (Montanarella et al. 2018). The IPBES report, together with this report focusing on climate
45 change, may contribute to create synergy between the two main global challenges for addressing land
46 degradation in order to help achieving the goals of SDG 15 (Protect, restore and promote sustainable
47 use of terrestrial ecosystems, sustainably manage forests, combat desertification, and halt and reverse
48 land degradation and halt biodiversity loss).

1 Market based mechanisms like the Clean Development Mechanism (CDM) under the UNFCCC and
2 the voluntary carbon market provide incentives to enhance carbon sinks on the land through
3 afforestation and reforestation. Implications for local land use and food security have been raised as a
4 concern and need to be assessed (Edstedt and Carton 2018b; Olsson et al. 2014b). Many projects
5 aimed at reducing emissions from deforestation and forest degradations (not to be confused with the
6 national REDD+ programs in accordance with the UNFCCC Warsaw Framework) are being planned
7 and implemented primarily targeting countries with high forest cover and high deforestation rates.
8 Some parameters of incentivising emissions reduction, quality of forest governance, conservation
9 priorities, local rights and tenure frameworks, and sub-national project potential are being looked into
10 with often very mixed results (Newton et al. 2016; Gebara and Agrawal 2017).

11 Besides international public initiatives, some actors in the private sector are increasingly aware of the
12 negative environmental impacts of some global value chains producing food, fibre, and energy
13 products (Lambin et al. 2018; van der Ven and Cashore 2018; van der Ven et al. 2018; Lyons-White
14 and Knight 2018). While improvement is under way in many supply chains, measures implemented so
15 far are often insufficient to be effective in reducing or stopping deforestation and forest degradation
16 (Lambin et al. 2018). The GEF is investing in actions to reduce deforestation in commodity supply
17 chains through its Food Systems, Land Use, and Restoration Impact Program (GEF 2018).

18 **4.8.5.1 Limits to adaptation**

19 SLM can be deployed as a powerful adaptation strategy in most instances of climate change impacts
20 on natural and social systems, yet there are limits to adaptation (Klein, R.J.T., G.F. Midgley, B.L.
21 Preston, M. Alam, F.G.H. Berkhout, K. Dow 2014; Dow et al. 2013a). Such limits are dynamic and
22 interact with social and institutional conditions (Barnett et al. 2015; Filho and Nalau 2018). Exceeding
23 adaptation limits will trigger escalating losses or require undesirable transformational change, such as
24 forced migration. The rate of change in relation to the rate of possible adaptation is crucial (Dow et al.
25 2013b). How limits to adaptation are defined and how they can be measured is contextual and
26 contested. Limits must be assessed in relation to the ultimate goals of adaptation, which is subject to
27 diverse and differential values (Dow et al. 2013b; Adger et al. 2009). A particularly sensitive issue is
28 whether migration is accepted as adaptation or not (Black et al. 2011; Tacoli 2009; Bardsley and
29 Hugo 2010). If migration were understood and accepted as a form of successful adaptation, it would
30 change the limits to adaptation by reducing or even avoiding future humanitarian crises caused by
31 climate extremes (Adger et al. 2009; Upadhyay et al. 2017; Nalau et al. 2018).

32 In the context of land degradation potential limits to adaptation exist if land degradation becomes so
33 severe and irreversible that livelihoods cannot be maintained, and if migration is either not acceptable
34 or possible. Examples are coastal erosion where land disappears (Gharbaoui and Blocher 2016; Luetz
35 2018), collapsing livelihoods due to thawing of permafrost (Landauer and Juhola 2019), and extreme
36 forms of soil erosion (e.g., landslides (Van der Geest and Schindler 2016) and gully erosion leading to
37 badlands (Poesen et al. 2003)).

38 **4.8.6 Resilience and thresholds**

39 Resilience refers to the capacity of interconnected social, economic and ecological systems, such as
40 farming systems, to absorb disturbance (e.g., drought, conflict, market collapse), and respond or
41 reorganise, to maintain their essential function, identity and structure. Resilience can be described as
42 “coping capacity”. The disturbance may be a shock - sudden events such as a flood or disease
43 epidemic – or it may be a trend that develops slowly, like a drought or market shift. The shocks and
44 trends anticipated to occur due to climate change are expected to exacerbate risk of land degradation.
45 Therefore, assessing and enhancing resilience to climate change is a critical component of designing
46 sustainable land management strategies.

1 Resilience as an analytical lens is particularly strong in ecology and related research on natural
2 resource management (Folke et al. 2010; Quinlan et al. 2016) while in the social sciences the
3 relevance of resilience for studying social and ecological interactions is contested (Cote and
4 Nightingale 2012; Olsson et al. 2015; Cretney 2014; Béné et al. 2012; Joseph 2013). In the case of
5 adaptation to climate change (and particularly regarding limits to adaptation), a crucial ambiguity of
6 resilience is the question whether resilience is a normative concept (i.e. resilience is good or bad) or is
7 a descriptive characteristic of a system (i.e. neither good nor bad). Previous IPCC reports have
8 defined resilience as a normative (positive) attribute (see AR5 Glossary), while the wider scientific
9 literature is divided on this (Weichselgartner and Kelman 2015; Strunz 2012; Brown 2014; Grimm
10 and Calabrese 2011; Thorén and Olsson 2018). For example, is outmigration from a disaster prone
11 area considered a successful adaptation (high resilience) or a collapse of the livelihood system (lack of
12 resilience) (Thorén and Olsson 2018)? In this report resilience is considered a positive attribute when
13 it maintains capacity for adaptation, learning and/or transformation.

14 Furthermore, resilience and the related terms adaptation and transformation are defined and used
15 differently by different communities (Quinlan et al. 2016). The relationship and hierarchy of
16 resilience with respect to vulnerability and adaptive capacity are also debated, with different
17 perspectives between the disaster management, and global change communities, (e.g., Cutter et al.
18 2008). Nevertheless, these differences in usage need not inhibit the application of “resilience
19 thinking” in managing land degradation; researchers using these terms, despite variation in
20 definitions, apply the same fundamental concepts to inform management of human-environment
21 systems, to maintain or improve the resource base, and sustain livelihoods.

22 Applying resilience concepts involves viewing the land as a component of an interlinked social-
23 ecological system; identifying key relationships that determine system function and vulnerabilities of
24 the system; identifying thresholds or tipping points beyond which the system transitions to an
25 undesirable state; and devising management strategies to steer away from thresholds of potential
26 concern, thus facilitating healthy systems and sustainable production (Walker et al., 2009).

27 A threshold is a non-linearity between a controlling variable and system function, such that a small
28 change in the variable causes the system to shift to an alternative state. Bestelmeyer et al. (2015) and
29 Prince et al. (2018) illustrate this concept in the context of land degradation. Studies have identified
30 various biophysical and socio-economic thresholds in different land-use systems. For example, 50%
31 ground cover (living and dead plant material and biological crusts) is a recognised threshold for
32 dryland grazing systems (e.g., Tighe et al. 2012); below this threshold infiltration rate declines, risk
33 of erosion causing loss of topsoil increases, a switch from perennial to annual grass species occurs and
34 there is a consequential sharp decline in productivity. This shift to a lower-productivity state cannot
35 be reversed without significant human intervention. Similarly, the combined pressure of water
36 limitations and frequent fire can lead to transition from closed forest to savannah or grassland: if fire
37 is too frequent trees do not reach reproductive maturity and post-fire regeneration will fail; likewise,
38 reduced rainfall / increased drought prevents successful forest regeneration (Reyer et al. 2015;
39 Thompson et al. 2009) see also Cross-chapter box 3 on Fire and climate change, Chapter 2.

40 In managing land degradation, it is important to assess the resilience of the existing system, and the
41 proposed management interventions. If the existing system is in an undesirable state or considered
42 unviable under expected climate trends, it may be desirable to promote adaptation or even
43 transformation to a different system that is more resilient to future changes. For example, in an
44 irrigation district where water shortages are predicted, measures could be implemented to improve
45 water use efficiency, for example by establishing drip irrigation systems for water delivery, although
46 transformation to pastoralism or mixed dryland cropping/livestock production may be more
47 sustainable in the longer term, at least for part of the area. Application of sustainable land
48 management practices, especially those focussed on ecological functions (e.g., agroecology,

ecosystem-based approaches, regenerative agriculture, organic farming), can be effective in building resilience of agro-ecosystems (Henry et al. 2018). Similarly, the resilience of managed forests can be enhanced by sustainable forest management that protects or enhances biodiversity, including assisted migration of tree species within their current range limit (Winder et al. 2011; Pedlar et al. 2012) or increasing species diversity in plantation forests (Felton et al. 2010; Liu et al. 2018a). The essential features of a resilience approach to management of land degradation under climate change are described by (O’Connell et al. 2016; Simonsen et al. 2014).

Consideration of resilience can enhance effectiveness of interventions to reduce or reverse land degradation (*medium agreement, limited evidence*). This approach will increase the likelihood that SLM/SFM and land restoration/rehabilitation interventions achieve long-term environmental and social benefits. Thus, consideration of resilience concepts can enhance the capacity of land systems to cope with climate change and resist land degradation, and assist land use systems to adapt to climate change.

4.8.7 Barriers to implementation of sustainable land management

There is a growing recognition that addressing barriers and designing solutions to complex environmental problems, such as land degradation, requires awareness of the larger system into which the problems and solutions are embedded (Laniak et al. 2013). An ecosystem approach to SLM based on understanding of the processes of land degradation has been recommended that can separate multiple drivers, pressures and impacts (Kassam et al. 2013), but large uncertainty in model projections of future climate, and associated ecosystem processes (IPCC 2013a) pose additional challenges to the implementation of SLM. As discussed earlier in this chapter, many SLM practices, including both technologies and approaches, are available that can increase yields and contribute to closing the yield gap between actual and potential crop or pasture yield, while also enhancing resilience to climate change (Yengoh and Ardö 2014; WOCAT). However, there are often systemic barriers to adoption and scaling up of SLM practices, especially in developing countries.

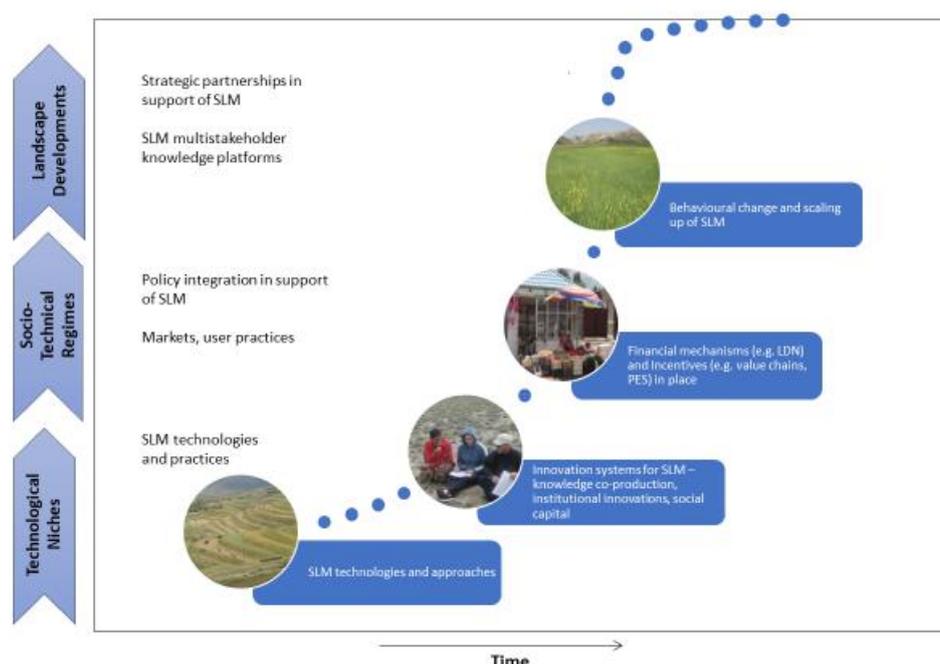
Uitto (2016) identified areas that the GEF, the financial mechanism of the UNCCD, UNFCCC and other multilateral environmental agreements, can address to solve global environmental problems. This includes removal of barriers related to knowledge and information; strategies for implementation of technologies and approaches; and institutional capacity. Strengthening these areas would drive transformational change leading to behavioral change and broader adoption of sustainable environmental practices. Detailed analysis of barriers as well as strategies, methods and approaches to scale up SLM have been undertaken for GEF programs in Africa, China and globally (Tengberg and Valencia 2018; Liniger et al. 2011; Tengberg et al. 2016). A number of interconnected barriers and bottlenecks to the scaling up of SLM have been identified in this context and are related to:

- Limited access to knowledge and information, including new SLM technologies and problem-solving capacities;
- Weak enabling environment, including the policy, institutional and legal framework for SLM, and land tenure and property rights;
- Inadequate learning and adaptive knowledge management in the project cycle, including monitoring and evaluation of impacts; and
- Limited access to finance for scaling up, including public and private funding, innovative business models for SLM technologies and financial mechanisms and incentives, such as payments for ecosystem services (PES), insurance and micro-credit schemes (see also Shames et al 2014).

Adoption of innovations and new technologies are increasingly analysed using the transition theory framework (Geels 2002), the starting point being the recognition that many global environmental problems cannot be solved by technological change alone but require more far-reaching change of

1 social-ecological systems. Using transition theory makes it possible to analyse how adoption and
2 implementation follow the four stages of sociotechnical transitions, from predevelopment of
3 technologies and approaches at the niche level, take-off and acceleration, to regime shift and
4 stabilisation at the landscape level. According to a recent review of sustainability transitions in
5 developing countries (Wieczorek 2018), three internal niche processes are important, including the
6 formation of networks that support and nurture innovation, the learning process and the articulation of
7 expectations to guide the learning process. While technologies are important, institutional and
8 political aspects form the major barriers to transition and upscaling. In developing and transition
9 economies, informal institutions play a pivotal role and transnational linkages are also important, such
10 as global value chains. In these countries, it is therefore more difficult to establish fully coherent
11 regimes or groups of individuals who share expectations, beliefs or behavior, as there is a high level
12 of uncertainty about rules and social networks or dominance of informal institutions, which creates
13 barriers to change. This uncertainty is further exacerbated by climate change. Landscape forces
14 comprise a set of slow changing factors, such as broad cultural and normative values, long-term
15 economic effects such as urbanisation, and shocks such as war and crises that can lead to change.

16 A study on SLM in the Kenyan highlands using transition theory concluded that barriers to adoption
17 of SLM included high poverty levels, a low input-low output farming system with limited potential to
18 generate income, diminishing land sizes and low involvement of the youth in farming activities.
19 Coupled with a poor coordination of government policies for agriculture and forestry, these barriers
20 created negative feedbacks in the SLM transition process. Other factors to consider include gender
21 issues and lack of secure land tenure. Scaling up of SLM technologies would require collaboration of
22 diverse stakeholders across multiple scales, a more supportive policy environment and substantial
23 resource mobilisation (Mutoko et al. 2014). Tengberg and Valencia (2018) analysed the findings from
24 a review of the GEF integrated natural resources management portfolio of projects using the transition
25 theory framework (Figure 4.7). They concluded that to remove barriers to SLM, an agricultural
26 innovations systems approach that supports co-production of knowledge with multiple stakeholders,
27 institutional innovations, a focus on value chains and strengthening of social capital to facilitate
28 shared learning and collaboration could accelerate the scaling up of sustainable technologies and
29 practices from the niche to the landscape level. Policy integration and establishment of financial
30 mechanisms and incentives could contribute to overcoming barriers to a regime shift. The new SLM
31 regime could in turn be stabilised and sustained at the landscape level by multi-stakeholder
32 knowledge platforms and strategic partnerships. However, transitions to more sustainable regimes and
33 practices are often challenged by lock-in mechanisms in the current system (Lawhon and Murphy
34 2012), such as economies of scale, investments already made in equipment, infrastructure and
35 competencies, lobbying, shared beliefs, and practices, which could hamper wider adoption of SLM.



1
2 **Figure 4.7 The transition from SLM niche adoption to regime shift and landscape development (figure**
3 **draws inspiration from (Geels 2002)). Adapted from (Tengberg and Valencia 2018)**

4 Adaptive, multi-level and participatory governance of social-ecological systems is considered
5 important for regime shifts and transitions to take place (Wieczorek 2018) and essential to secure the
6 capacity of environmental assets to support societal development over longer time periods (Folke et
7 al. 2005). There is also recognition that effective environmental policies and programs need to be
8 informed by a comprehensive understanding of the biophysical, social, and economic components and
9 processes of a system, their complex interactions, and how they respond to different changes (Kelly
10 (Letcher) et al. 2013). But blueprint policies will not work due to the wide diversity of rules and
11 informal institutions used across sectors and regions of the world, especially in traditional societies
12 (Ostrom 2009).

13 The most effective way of removing barriers to funding of SLM has been mainstreaming of SLM
14 objectives and priorities into relevant policy and development frameworks and combining SLM best
15 practices with economic incentives for land users. As the short-term costs for establishing and
16 maintaining SLM measures are generally high and constitute a barrier to adoption, land users may
17 need to be compensated for generation of longer-term public goods, such as ecosystem services. Cost-
18 benefit analyses can be conducted on SLM interventions to facilitate such compensations (Liniger et
19 al. 2011; Nkonya et al. 2016; Tengberg et al. 2016). The landscape approach is a means to reconcile
20 competing demands on the land and remove barriers to implementation of SLM (e.g. Sayer et al.
21 2013; Bürgi et al. 2017). It involves an increased focus on participatory governance, development of
22 new SLM business models, and innovative funding schemes including insurance (Shames et al. 2014).
23 The Land Degradation Neutrality (LDN) Fund takes a landscape approach and raises private finance
24 for SLM and promotes market-based instruments, such as Payment for Ecosystem Services (PES),
25 certification and carbon trading, that can support scaling up of SLM to improve local livelihoods,
26 sequester carbon and enhance the resilience to climate change.

27 4.9 Case-studies

28 Climate change impacts on land degradation can be avoided, reduced or even reversed, but need to be
29 addressed in a context sensitive manner. Many of the responses described in this section can also

1 provide synergies of adaptation and mitigation. In this section we provide more in-depth analysis of a
 2 number of salient aspects of how land degradation and climate change interact. Table 4.3 is a
 3 synthesis of how of these case studies relate to climate change and other broader issues in terms of co-
 4 benefits.

5

6 **Table 4.3 Synthesis of how the case studies interact with climate change and a broader set of co-benefits**

| Case studies (4.10) | (4.9) | Mitigation benefits and potential | Adaptation benefits | Co-benefits | Legend | |
|---|---------|---|---|--|---|-----------------------------------|
| Urban green infrastructure (4.10.1) An increasing majority of the world population live in cities and land degradation is an urgent matter for urban areas | (4.9.1) |  |  | human health, recreation |  | carbon sink |
| Perennial grains (4.10.2) After 40 years of breeding, perennial grains now seem to have the potential of reducing climate impacts of agriculture while increasing its overall sustainability | (4.9.2) |  |  | reduced use of herbicides, reduced soil erosion and nutrient leakage |  | reduced emission |
| Reforestation (4.10.3) Two cases of successful reforestation serve as illustrations of the potential of sustained efforts into reforestation | (4.9.3) |  |  | economic return from sustainable forestry, reduced flood risk downstream | | |
| Management of peat soils (4.10.4) Degradation of peat soils in tropical and arctic regions is a major source of greenhouse gases, hence an urgent mitigation option | (4.9.4) |  | | improved air quality in tropical regions |  | reduced flood risk |
| Biochar (4.10.5) Biochar is a land management technique of high potential, but controversial | (4.9.5) |  | | improved soil fertility |  | reduced heat stress |
| Protection against hurricane damages (4.10.6) More severe tropical cyclones increase the risk of land degradation in some areas, hence the need for increased adaptation | (4.9.6) | |  | reduced losses (human lives, livelihoods, and assets) |  | drought resistance |
| Responses to salt water intrusion (4.10.7) The combined effect of climate induced sea level rise and land use change in coastal regions increases the risk of saltwater intrusion in many coastal regions | (4.9.7) | |  | improved food and water security, |  | storm protection |
| Avoiding coastal maladaptation (4.10.8) Low lying coastal areas are in urgent need of adaptation, but examples have resulted in maladaptation | (4.9.8) | |  | reduced losses (human lives, livelihoods, and assets) |  | protection against sea level rise |

7

8 **4.9.1 Urban green infrastructure**

9 Over half the world's population now lives in towns and cities, a proportion that is predicted to
 10 increase to ~70% by the middle of the century (United Nations 2015). Rapid urbanisation is a severe
 11 threat to land and the provision of ecosystem services (Seto et al. 2012). However, as cities expand,
 12 the avoidance of land degradation, or the maintenance/enhancement of ecosystem services is rarely
 13 considered in planning processes. Instead economic development and the need for space for
 14 construction is prioritised, which can result in substantial pollution of air and water sources, the
 15 degradation of existing agricultural areas and indigenous, natural or semi-natural ecosystems both
 16 within and outside of urban areas. For instance, urban areas are characterised by extensive impervious
 17 surfaces. Degraded, sealed soils beneath these surfaces do not provide the same quality of water
 18 retention as intact soils. Urban landscapes comprising 50-90% impervious surfaces can therefore
 19 result in 40-83% of rainfall becoming surface water runoff (Pataki et al. 2011). With rainfall intensity
 20 predicted to increase in many parts of the world under climate change (Royal Society 2016), increased

1 water runoff is going to get worse. Urbanisation, land degradation and climate change are therefore
2 strongly interlinked, suggesting the need for common solutions (Reed and Stringer 2016b).

3 There is now a large body of research and application demonstrating the importance of retaining
4 urban green infrastructure (UGI) for the delivery of multiple ecosystem services (DG Environment
5 News Alert Service, 2012; Wentworth, 2017) as an important tool to mitigate and adapt to climate
6 change. UGI can be defined as all green elements within a city, including but not limited to retained
7 indigenous ecosystems, parks, public greenspaces, green corridors, street trees, urban forests, urban
8 agriculture, green roofs/walls and private domestic gardens (Tzoulas et al. 2007). The definition is
9 usually extended to include ‘blue’ infrastructure, such as rivers, lakes, bioswales and other water
10 drainage features. The related concept of Nature Based Solutions (defined as: living solutions inspired
11 by, continuously supported by and using nature, which are designed to address various societal
12 challenges in a resource-efficient and adaptable manner and to provide simultaneously economic,
13 social, and environmental benefits) has gained considerable traction within the European Commission
14 as one approach to mainstreaming the importance of UGI (Maes and Jacobs 2017; European Union
15 2015).

16 Through retaining existing vegetation and ecosystems, revegetating previous developed land or
17 integrating vegetation into buildings in the form of green walls and roofs, UGI can play a direct role
18 in mitigating climate change through carbon sequestration. However, compared to overall carbon
19 emissions from cities, effects will be small. Given that UGI necessarily involves the retention and
20 management of non-sealed surfaces, co-benefits for land degradation (e.g. soil compaction avoidance,
21 reduced water run-off, carbon storage and vegetation productivity; (Davies et al. 2011; Edmondson et
22 al. 2011, 2014; Yao et al. 2015) will also be apparent. Although not currently a priority, its role in
23 mitigating land degradation could be substantial. For instance, appropriately managed innovative
24 urban agricultural production systems, such as vertical farms, could have the potential to both meet
25 some of the food needs of cities and reduce the production (and therefore degradation) pressure on
26 agricultural land in rural areas, although thus far this is unproven (for a recent review (Wilhelm and
27 Smith 2018)).

28 The importance of UGI as part of a climate change adaptation approach has received greater attention
29 and application (Gill et al. 2007; Fryd et al. 2011; Demuzere et al. 2014; Sussams et al. 2015). The
30 EU’s Adapting to Climate Change White Paper emphasises the “crucial role in adaptation in
31 providing essential resources for social and economic purposes under extreme climate conditions”
32 (CEC, 2009, p. 9). Increasing vegetation cover, planting street trees and maintaining/expanding public
33 parks reduces temperatures (Cavan et al. 2014; Di Leo et al. 2016; Feyisa et al. 2014; Tonosaki K,
34 Kawai S 2014; Zölch et al. 2016). Further, the appropriate design and spatial distribution of
35 greenspaces within cities can help to alter urban climates to improve human health and comfort (e.g.
36 (Brown and Nicholls 2015; Klemm et al. 2015)). The use of green walls and roofs can also reduce
37 energy use in buildings (e.g. (Coma et al. 2017)). Similarly, natural flood management and ecosystem
38 based approaches of providing space for water, renaturalising rivers and reducing surface run-off
39 through the presence of permeable surfaces and vegetated features (including walls and roofs) can
40 manage flood risks, impacts and vulnerability (e.g. (Gill et al. 2007; Munang et al. 2013)). Access to
41 UGI in times of environmental stresses and shock can provide safety nets for people and can,
42 therefore, be an important adaptation mechanism, both to climate change (Potschin et al. 2016) and
43 land degradation.

44 Most examples of UGI implementation as a climate change adaptation strategy have centered on its
45 role in water management for flood risk reduction. The importance for land degradation is either not
46 stated, or not prioritized. In Beira, Mozambique, the government is using UGI to mitigate against
47 increased flood risks predicted to occur under climate change and urbanisation, which will be done by
48 improving the natural water capacity of the Chiveve River. As part of the UGI approach, mangrove

1 habitats have been restored and future phases include developing new multi-functional urban green
2 spaces along the river (World Bank 2016). The retention of green spaces within the city will have the
3 added benefit of halting further degradation in those areas. Elsewhere, planning mechanisms promote
4 the retention and expansion of green areas within cities to ensure ecosystem service delivery, which
5 directly halts land degradation, but are largely viewed and justified in the context of climate change
6 adaptation and mitigation. For instance, the Landscape Programme in Berlin includes five plans, one
7 of which covers adapting to climate change through the recognition of the role of UGI (Green Surge
8 2016). Major climate related challenges facing Durban, South Africa, include sea level rise, urban
9 heat island, water runoff and conservation (Roberts and O'Donoghue 2013). Now considered a global
10 leader in climate adaptation planning (Roberts 2010), Durban's Climate Change Adaptation plan
11 includes the retention and maintenance of natural ecosystems in particular those which are important
12 for mitigating flooding, coastal erosion, water pollution, wetland siltation and climate change
13 (eThekweni Municipal Council 2014).

14 **4.9.2 Perennial Grains and Soil Organic Carbon**

15 The severe ecological perturbation that is inherent in the conversion of native perennial vegetation to
16 annual crops, and the subsequent high frequency of perturbation required to maintain annual crops,
17 results in at least four forms of soil degradation that will be exacerbated by the effects of climate
18 change (Crews et al. 2016). First, soil erosion is a very serious consequence of annual cropping with
19 median losses exceeding rates of formation by 1-2 orders of magnitude in conventionally plowed
20 agroecosystems, and while erosion is reduced with conservation tillage, median losses still exceed
21 formation by several fold (Montgomery 2007). More severe storm intensity associated with climate
22 change is expected to cause even greater losses to wind and water erosion (Nearing et al. 2004b).
23 Secondly, the periods of time in which live roots are reduced or altogether absent from soils in annual
24 cropping systems allow for substantial losses of nitrogen from fertilised croplands, averaging 50%
25 globally (Ladha et al. 2005). This low retention of nitrogen is also expected to worsen with more
26 intense weather events (Bowles et al. 2018). A third impact of annual cropping is the degradation of
27 soil structure caused by tillage, which can reduce infiltration of precipitation, and increase surface
28 runoff. It is predicted that the percentage of precipitation that infiltrates into agricultural soils will
29 decrease further under climate change scenarios (Basche and DeLonge 2017; Wuest et al. 2006). The
30 fourth form of soil degradation that results from annual cropping is the reduction of soil organic
31 matter (SOM), a topic of particular relevance to climate change mitigation and adaptation.

32 Undegraded cropland soils can theoretically hold far more SOM (which is ~58% carbon) than they
33 currently do (Soussana et al. 2006). We know this deficiency because, with few exceptions,
34 comparisons between cropland soils and those of proximate mature native ecosystems commonly
35 show a 40-75% decline in soil carbon attributable to agricultural practices. What happens when native
36 ecosystems are converted to agriculture that induces such significant losses of SOM? Wind and water
37 erosion commonly results in preferential removal of light organic matter fractions that can accumulate
38 on or near the soil surface (Lal 2003). In addition to the effects of erosion, the fundamental practices
39 of growing annual food and fiber crops alters both inputs and outputs of organic matter from most
40 agroecosystems resulting in net reductions in soil carbon equilibria (Soussana et al. 2006;
41 McLauchlan 2006; Crews et al. 2016). Native vegetation of almost all terrestrial ecosystems is
42 dominated by perennial plants, and the belowground carbon allocation of these perennials is a key
43 variable in determining formation rates of stable soil organic carbon (SOC) (Jastrow et al. 2007;
44 Schmidt et al. 2011). When perennial vegetation is replaced by annual crops, inputs of root-associated
45 carbon (roots, exudates, mycorrhizae) decline substantially. For example, perennial grassland species
46 allocate around 67% of productivity to roots, whereas annual crops allocate between 13-30% (Saugier
47 2001; Johnson et al. 2006).

1 At the same time inputs of SOC are reduced in annual cropping systems, losses are increased because
2 of tillage, compared to native perennial vegetation. Tillage breaks apart soil aggregates, which, among
3 other functions, are thought to inhibit soil bacteria, fungi and other microbes from consuming and
4 decomposing soil organic matter (Grandy and Neff 2008). Aggregates reduce microbial access to
5 organic matter by restricting physical access to mineral-stabilized organic compounds as well as
6 reducing oxygen availability (Cotrufo et al. 2015; Lehmann and Kleber 2015). When soil aggregates
7 are broken open with tillage in the conversion of native ecosystems to agriculture, microbial
8 consumption of SOC and subsequent respiration of CO₂ increase dramatically, reducing soil carbon
9 stocks (Grandy and Robertson 2006; Grandy and Neff 2008).

10 Many management approaches are being evaluated to reduce soil degradation in general, especially
11 by increasing mineral-protected forms of SOC in the world's croplands (Paustian et al. 2016). The
12 menu of approaches being investigated focus either on increasing belowground carbon inputs, usually
13 through increases in total crop productivity, or by decreasing microbial activity, usually through
14 reduced soil disturbance (Crews and Rumsey 2017). However, the basic biogeochemistry of terrestrial
15 ecosystems managed for production of annual crops presents serious challenges to achieving the
16 standing stocks of SOC accumulated by native ecosystems that preceded agriculture. A novel new
17 approach that is just starting to receive significant attention is the development of perennial cereal,
18 legume and oilseed crops (Glover et al. 2010; Baker 2017).

19 There are two basic strategies that plant breeders and geneticists are using to develop new perennial
20 grain crop species. The first involves making wide hybrid crosses between existing elite lines of
21 annual crops, such as wheat, sorghum and rice, with related wild perennial species in order to
22 introgress perennialism into the genome of the annual (Cox et al. 2018; Huang et al. 2018; Hayes et
23 al. 2018). The other approach is *de novo* domestication of wild perennial species that have crop-like
24 traits of interest (DeHaan et al. 2016; DeHaan and Van Tassel 2014). New perennial crop species
25 undergoing *de novo* domestication include intermediate wheatgrass, a relative of wheat that produces
26 grain also known as Kernza (DeHaan et al. 2018; Cattani and Asselin 2018) and *Silphium*
27 *integrifolium*, an oilseed crop in the sunflower family (Van Tassel et al. 2017). Other perennial grain
28 crops receiving attention include pigeon pea, barley, buckwheat and maize (Batello et al. 2014; Chen
29 et al. 2018c) and a number of legume species (Schlautman et al. 2018). In most cases, the seed yields
30 of perennial grain crops under development are well below those of elite modern grain varieties. In
31 the time that it takes intensive breeding efforts to close the yield and other trait gaps between annual
32 and perennial grains, perennial proto-crops may be used for purposes other than grain, including
33 forage production (Ryan et al. 2018). Perennial rice stands out as a high-yielding exception, as its
34 yields matched those of elite local varieties in the Yunnan Province for six growing seasons over three
35 years (Huang et al. 2018).

36 In a perennial agroecosystem, the biogeochemical controls on SOC accumulation shift dramatically,
37 and begin to resemble the controls that govern native ecosystems (Crews et al. 2016). When erosion
38 is reduced or halted, and crop allocation to roots increases by 100-200%, and when soil aggregates are
39 not disturbed thus reducing microbial respiration, SOC levels are expected to increase (Crews and
40 Rumsey 2017). Deep roots growing year-round are also effective at increasing nitrogen retention
41 (Culman et al. 2013; Jungers et al. 2019). Substantial increases in SOC have been measured where
42 croplands that had historically been planted to annual grains were converted to perennial grasses, such
43 as in the Conservation Reserve Program (CRP) of the US, or in plantings of second generation
44 perennial biofuel crops. Two studies have assessed carbon accumulation in soils when croplands were
45 converted to the perennial grain Kernza. In one, researchers found no differences in soil labile
46 (permanganate-oxidizable) C after 4 years of cropping to perennial Kernza versus annual wheat in a
47 sandy textured soil. Given that coarse textured soils do not offer the same physicochemical protection
48 against microbial attack as many finer textured soils, these results are not surprising, but these results

1 do underscore how variable rates of carbon accumulation can be (Jastrow et al. 2007). In the second
2 study, researchers assessed the carbon balance of a Kernza field in Kansas USA over 4.5 years using
3 eddy covariance observations (de Oliveira et al. 2018). They found the net C accumulation rate of
4 about $1500 \text{ g C m}^{-2} \text{ yr}^{-1}$ in the first year of the study corresponding to the biomass of Kernza
5 increasing, to about $300 \text{ g C m}^{-2} \text{ yr}^{-1}$ in the final year where CO_2 respiration losses from the
6 decomposition of roots and soil organic matter approached new carbon inputs from photosynthesis.
7 Based on measurements of soil carbon accumulation in restored grasslands in this part of US, the net
8 carbon accumulation in stable organic matter under a perennial grain crop might be expected to
9 sequester $30\text{-}50 \text{ g C m}^{-2} \text{ yr}^{-1}$ (Post and Kwon 2000) until a new equilibrium is reached. Sugar cane, a
10 highly productive perennial, has been shown to accumulate a mean of $187 \text{ g C m}^{-2} \text{ yr}^{-1}$ in Brazil (La
11 Scala Júnior et al. 2012).

12 Reduced soil erosion, increased nitrogen retention, greater water uptake efficiency and enhanced
13 carbon sequestration represent improved ecosystem functions made possible in part by deep and
14 extensive root systems of perennial crops (Figure 4.8).



15
16 **Figure 4.8 Comparison of root systems between the newly domesticated intermediate wheatgrass (left)**
17 **and annual wheat (right). Photo and copyright: Jim Richardson**

18 When compared to annual grains like wheat, single species stands of deep rooted perennial grains
19 such as Kernza are expected to reduce soil erosion, increase nitrogen retention, achieve greater water
20 uptake efficiency and enhance carbon sequestration (Crews et al. 2018) (Figure 4.8). An even higher
21 degree of ecosystem services can at least theoretically be achieved by strategically combining
22 different functional groups of crops such as a cereal and a nitrogen-fixing legume (Soussana and
23 Lemaire 2014). Not only is there evidence from plant diversity experiments that communities with
24 higher species richness sustain higher concentrations of soil organic carbon (Hungate et al. 2017;
25 Sprunger and Robertson 2018; Chen et al. 2018b; Yang et al. 2019), but other valuable ecosystem
26 services such as pest suppression, lower greenhouse gas emissions, and greater nutrient retention may
27 be enhanced (Schnitzer et al. 2011; Culman et al. 2013).

1 Similar to perennial forage crops such as alfalfa, perennial grain crops are expected to have a definite
2 productive life span, probably in the range of 3-10 years. A key area of research on perennial grains
3 cropping systems is to minimise losses of soil organic carbon during conversion of one stand of
4 perennial grains to another. Recent work demonstrates that no-till conversion of a mature perennial
5 grassland to another perennial crop will experience several years of high net CO₂ emissions as
6 decomposition of copious crop residues exceeds ecosystem uptake of carbon by the new crop (Abraha
7 et al. 2018). Most if not all of this lost carbon will be recaptured in the replacement crop. It is not
8 known whether mineral-stabilised carbon that is protected in soil aggregates is vulnerable to loss in
9 perennial crop succession.

10 Perennial grains hold promises of agricultural practices which can significantly reduce soil erosion
11 and nutrient leakage while sequestering carbon. When cultivated in mixes with N-fixing species
12 (legumes) such polycultures also reduce the need for external inputs of nitrogen - a large source of
13 GHG from conventional agriculture.

14 **4.9.3 Reversing land degradation through reforestation**

15 **4.9.3.1 South Korea Case Study on Reforestation Success**

16 In the first half of the 20th century, forests in the Republic of South Korea were severely degraded and
17 deforested during foreign occupations and the Korean War. Unsustainable harvest for timber and fuel
18 wood resulted in severely degraded landscapes, heavy soil erosion and large areas denuded of
19 vegetation cover. Recognising that South Korea's economic health would depend on a healthy
20 environment, South Korea established a national forest service (1967) and embarked on the first phase
21 of a 10-year reforestation program in 1973 (Forest Development Program), which was followed by
22 subsequent reforestation programs that ended in 1987, after 2.4 Mha of forests were restored, see
23 Figure 4.9.

24 As a consequence of reforestation, forest volume increased from 11.3 m³ ha⁻¹ in 1973 to 125.6 m³ ha⁻¹
25 in 2010 and 150.2 m³ ha⁻¹ in 2016 (Korea Forest Service 2017). Increases in forest volume had
26 significant co-benefits such as increasing water yield by 43% and reducing soil losses by 87% from
27 1971 to 2010 (Kim et al. 2017).

28 The forest carbon density in South Korea has increased from 5–7 Mg C ha⁻¹ in the period 1955–1973
29 to more than 30 Mg C ha⁻¹ in the late 1990s (Choi et al. 2002). Estimates of C uptake rates in the late
30 1990s were 12 Tg C yr⁻¹ (Choi et al. 2002). For the period 1954 to 2012 C uptake was 8.3 Tg C yr⁻¹
31 (Lee et al. 2014), lower than other estimates because reforestation programs did not start until 1973.
32 NEP in South Korea was 10.55 ± 1.09 Tg C yr⁻¹ in the 1980s, 10.47 ± 7.28 Tg C yr⁻¹ in the 1990s,
33 and 6.32 ± 5.02 Tg C yr⁻¹ in the 2000s, showing a gradual decline as average forest age increased
34 (Cui et al. 2014). The estimated past and projected future increase in the carbon content of South
35 Korea's forest area during 1992-2034 was 11.8 Tg C yr⁻¹ (Kim et al. 2016).



1

2 **Figure 4.9 Example of severely degraded hills in South Korea and stages of forest restoration. The top**
 3 **two photos are taken in the early 1970s, before and after restoration, the third photo about 5 years after**
 4 **restoration and the bottom photo was taken about 20 years after restoration. Many examples of such**
 5 **restoration success exist throughout South Korea (Source: Korea Forest Service).**

6 During the period of forest restoration, South Korea also promoted inter-agency cooperation and
 7 coordination, especially between the energy and forest sectors, to replace firewood with fossil fuels,
 8 and by reducing demand for firewood helped forest recovery (Bae et al. 2012). As experience with
 9 forest restoration programs has increased, emphasis has shifted from fuelwood plantations, often with
 10 exotic species and hybrid varieties to planting more native species and encouraging natural
 11 regeneration (Kim and Zsuffa 1994; Lee et al. 2015). Avoiding monocultures in reforestation
 12 programs can reduce susceptibility to pests (Kim and Zsuffa 1994). Other important factors in the
 13 success of the reforestation program were that private landowners were heavily involved in initial
 14 efforts (both corporate entities and smallholders) and that the reforestation program was made part of
 15 the national economic development program (Lamb 2014).

16 The net present value and the benefit-cost ratio of the reforestation program were USD 54.3 billion
 17 and 5.84 billion in 2010, respectively. The breakeven point of the reforestation investment appeared
 18 within two decades. Substantial benefits of the reforestation program included disaster risk reduction
 19 and carbon sequestration (Lee et al. 2018a).

20 In summary, the reforestation program was a comprehensive technical and social initiative that
 21 restored forest ecosystems, enhanced the economic performance of rural regions, contributed to
 22 disaster risk reduction, and enhanced carbon sequestration (Kim et al. 2017; Lee et al. 2018a; UNDP
 23 2017).

1 The success of the reforestation program in South Korea and the associated significant carbon sink
2 indicate a high mitigation potential that might be contributed by a potential future reforestation
3 program in the Democratic People’s Republic of Korea (North Korea) (Lee et al. 2018b).

4 **4.9.3.2 China Case Study on Reforestation Success**

5 The dramatic decline in the quantity and quality of natural forests in China resulted in land
6 degradation, such as soil erosion, floods, droughts, carbon emission, and damage to wildlife habitat
7 (Liu and Diamond 2008). In response to failures of previous forestry and land policies, the severe
8 droughts in 1997, and the massive floods in 1998, the central government decided to implement a
9 series of land degradation control policies, including the National Forest Protection Program (NFPP),
10 Grain for Green or the Conversion of Cropland to Forests and Grasslands Program (GFGP) (Liu et al.
11 2008; Yin 2009; Tengberg et al. 2016; Zhang et al. 2000). The NFPP aimed to completely ban
12 logging of natural forests in the upper reaches of the Yangtze and Yellow rivers as well as in Hainan
13 Province by 2000 and to substantially reduce logging in other places (Xu et al. 2006). In 2011, NFPP
14 was renewed for the 10-year second phase, which also added another 11 counties around Danjiangkou
15 Reservoir in Hubei and Henan Provinces, the water source for the middle route of the South-to-North
16 Water Diversion Project (Liu et al. 2013). Furthermore, the NFPP afforested 31 Mha by 2010 through
17 aerial seeding, artificial planting, and mountain closure (i.e., prohibition of human activities such as
18 fuelwood collection and livestock grazing) (Xu et al. 2006). China banned commercial logging in all
19 natural forests by the end of 2016, which imposed logging bans and harvesting reductions in 68.2
20 Mha of forest land – including 56.4 Mha of natural forest (approximately 53% of China’s total natural
21 forests).

22 GFGP became the most ambitious of China’s ecological restoration efforts with over USD 45 billion
23 devoted to its implementation since 1990 (Kolinjivadi and Sunderland 2012) The program involves
24 the conversion of farmland on slopes of 15-25° or greater to forest or grassland (Bennett 2008). The
25 pilot program started in three provinces –Sichuan, Shaanxi, and Gansu – in 1999 (Liu and Diamond
26 2008). After initial success, it was extended to 17 provinces by 2000 and finally to all provinces by
27 2002, including the headwaters of the Yangtze and Yellow rivers (Liu et al. 2008).

28 NFPP and GFGP have dramatically improved China’s land conditions and ecosystem services, and
29 thus have mitigated the unprecedented land degradation in China (Liu et al. 2013; Liu et al 2002;
30 Long et al. 2006; Xu et al. 2006). NFPP protected 107 Mha forest area and increased forest area by 10
31 Mha between 2000 and 2010. For the second phase (2011–2020), the NFPP plans to increase forest
32 cover by a further 5.2 Mha, capture 416 million tons of carbon, provide 648,500 forestry jobs, further
33 reduce land degradation, and enhance biodiversity (Liu et al. 2013). During 2000–2007, sediment
34 concentration in the Yellow River had declined by 38%. In the Yellow River basin, it was estimated
35 that surface runoff would be reduced by 450 million m³ from 2000 to 2020, which is equivalent to
36 0.76% of the total surface water resources (Jia et al. 2006). GFGP had cumulatively increased
37 vegetative cover by 25 Mha, with 8.8 Mha of cropland being converted to forest and grassland, 14.3
38 Mha barren land being afforested, and 2.0 million ha of forest regeneration from mountain closure.
39 Forest cover within the GFGP region has increased 2% during the first 8 years (Liu et al. 2008). In
40 Guizhou Province, GFGP plots had 35–53% less loss of phosphorus than non-GFGP plots (Liu et al.
41 2002). In Wuqi County of Shaanxi Province, the Chaigou Watershed had 48% and 55% higher soil
42 moisture and moisture-holding capacity in GFGP plots than in non-GFGP plots, respectively (Liu et
43 al. 2002). According to reports on China’s first national ecosystem assessment (2000–2010), for
44 carbon sequestration and soil retention, coefficients for the GTGP targeting forest restoration and
45 NFPP are positive and statistically significant. For sand fixation, GTGP targeting grassland
46 restoration is positive and statistically significant. Remote sensing observations confirm vegetation
47 cover increases and bare soil decline in China over the period 2001 to 2015 (Qiu et al. 2017) (Qiu et
48 al. 2017). But where afforestation is sustained by drip irrigation from groundwater, questions about

1 plantation sustainability arise (Chen et al. 2018a). Moreover, greater gains in biodiversity could be
2 achieved by promoting mixed forests over monocultures (Hua et al. 2016).

3 NFPP-related activities received a total commitment of 93.7 billion yuan (about USD 14 billion with
4 today's exchange rate) between 1998 and 2009. Most of the money was used to offset economic
5 losses of forest enterprises caused by the transformation from logging to tree plantations and forest
6 management (Liu et al. 2008). By 2009, the cumulative total investment through the NFPP and GFGP
7 exceeded USD 50 billion and directly involved more than 120 million farmers in 32 million
8 households in the GFGP alone (Liu et al. 2013). All programs reduce or reverse land degradation and
9 improve human well-being. Thus, a coupled human and natural systems perspective (Liu et al. 2008)
10 would be helpful to understand the complexity of policies and their impacts, and to establish long-
11 term management mechanisms to improve the livelihood of participants in these programs and other
12 land management policies in both China and many other parts of the world.

13 **4.9.4 Degradation and management of peat soils**

14 Globally, peatlands cover 3-4 % of the Earth's land area (~430 Mha) (Xu et al. 2018a; Wu et al.
15 2017b) and store 26-44% of estimated global soil organic carbon (Moore 2002). They are most
16 abundant in high northern latitudes, covering large areas in North America, Russia and Europe. At
17 lower latitudes, the largest areas of tropical peatlands are located in Indonesia, the Congo Basin and
18 the Amazon Basin in the form of peat swamp forests (Gumbrecht et al. 2017; Xu et al. 2018a). It is
19 estimated that while 80-85% of the global peatland areas is still largely in a natural state, they are such
20 carbon-dense ecosystems that degraded peatlands (0.3% of the terrestrial land) are responsible for a
21 disproportional 5% of global anthropogenic carbon dioxide (CO₂) emissions, that is an annual
22 addition of 0.9-3 Gt of CO₂ to the atmosphere (Dommain et al. 2012; IPCC 2014c).

23 Peatland degradation is not well quantified globally, but regionally peatland degradation can involve a
24 large percentage of the areas. Land-use change and degradation in tropical peatlands have primarily
25 been quantified in Southeast Asia, where drainage and conversion to plantation crops is the dominant
26 transition (Miettinen et al. 2016). Degradation of peat swamps in Peru is also a growing concern and
27 one pilot survey showed that over 70% of the peat swamps were degraded in one region that was
28 surveyed (Hergoualc'h et al. 2017a). Around 65,000km² or 10% of the European peatland area has
29 been lost and 44% of the remaining European peatlands are degraded (Joosten, H., Tanneberger
30 2017). Large areas of fens have been entirely 'lost' or greatly reduced in thickness due to peat
31 wastage (Lamers et al. 2015).

32 The main drivers of the acceleration of peatland degradation in the twentieth century were associated
33 with drainage for agriculture, peat extraction and afforestation related activities (burning, over-
34 grazing, fertilisation) with a variable scale and severity of impact depending on existing resources in
35 the various countries (O'Driscoll et al. 2018; Abu et al. 2017; Dommain et al. 2018; Lamers et al.
36 2015). New drivers include urban development, wind farm construction (Smith et al. 2012), hydro-
37 electric development, tar sands mining and recreational (Joosten, H., Tanneberger 2017).
38 Anthropogenic pressures are now affecting peatlands in previously geographically isolated areas with
39 consequences for global environmental concerns and impacts on local livelihoods (Dargie et al. 2017;
40 Lawson et al. 2015; Butler et al. 2009).

41 Drained and managed peatlands are GHG emissions hotspots (Swails et al. 2018; Hergoualc'h et al.
42 2017b; Roman-Cuesta et al. 2016; Hergoualc'h et al. 2017a). In most cases, lowering of the water
43 table leads to direct and indirect CO₂ and N₂O emissions to the atmosphere with rates dependent on a
44 range of factors, including the groundwater level and the water content of surface peat layers, nutrient
45 content, temperature, and vegetation communities. The exception is nutrient limited boreal peatlands
46 (Minkinen et al. 2018; Ojanen et al. 2014). Drainage also increases erosion and dissolved organic C

1 loss, removing stored carbon into streams as dissolved and particulate organic carbon, which
2 ultimately returns to the atmosphere (Moore et al. 2013; Evans et al. 2016).

3 In tropical peatlands, oil palm is the most widespread plantation crop and on average it emits around
4 40 t CO₂ ha⁻¹ yr⁻¹; Acacia plantations for pulpwood are the second most widespread plantation crop
5 and emit around 73 t CO₂ ha⁻¹ yr⁻¹ (Drösler et al. 2013). Other land uses typically emit less than 37 t
6 CO₂ ha⁻¹ yr⁻¹. Total emissions from peatland drainage in the region are estimated to be between 0.07
7 and 1.1 Gt CO₂ yr⁻¹ (Houghton and Nassikas 2017; Frohking et al. 2011). Land-use change also affects
8 the fluxes of N₂O and CH₄. Undisturbed tropical peatlands emit about 0.8 Mt CH₄ yr⁻¹ and 0.002 Mt
9 N₂O yr⁻¹, while disturbed peatlands emit 0.1 Mt CH₄ yr⁻¹ and 0.2 Mt N₂O–N yr⁻¹ (Frohking et al. 2011).
10 These N₂O emissions are probably low as new findings show that emissions from fertilised oil palm
11 can exceed 20 kg N₂O–N ha⁻¹ yr⁻¹ (Oktarita et al. 2017).

12 In the temperate and boreal zones, peatland drainage often leads to emissions on the order of 0.9 to
13 9.5 t CO₂ ha⁻¹ y⁻¹ in forestry plantations and 21 to 29 t CO₂ ha⁻¹ y⁻¹ in grasslands and croplands.
14 Nutrient poor sites often continue to be CO₂ sinks for long periods (e.g. 50 y) following drainage and
15 in some cases sinks for atmospheric CH₄, even when drainage ditch emissions are considered
16 (Minkinen et al. 2018; Ojanen et al. 2014). Undisturbed boreal and temperate peatlands emit about
17 0.30 Mt CH₄ yr⁻¹ and 0.02 Mt N₂O–N yr⁻¹, while disturbed peatlands emit 0.1 Mt CH₄ yr⁻¹ and 0.2 Mt yr⁻¹
18 ¹N₂O (Frohking et al. 2011).

19 Fire emissions from tropical peatlands are only a serious issue in Southeast Asia, where they are
20 responsible for 634 (66–4070) Mt CO₂ yr⁻¹ (van der Werf et al. 2017). Much of the variability is
21 linked with the El Niño Southern Oscillation, which produces drought conditions in this region.
22 Anomalously active fire seasons have also been observed in non-drought years and this has been
23 attributed to the increasing effect of high temperatures that dry vegetation out during short dry spells
24 in otherwise normal rainfall years (Fernandes et al. 2017; Gaveau et al. 2014). Fires have significant
25 societal impacts; for example, the 2015 fires caused over 100,000 additional deaths across Indonesia,
26 Malaysia and Singapore and this event was more than twice as deadly as the 2006 El Niño event
27 (Koplitz et al. 2016).

28 Peatland degradation in other parts of the world differs from Asia. In Africa large peat deposits like
29 those found in the Cuvette Centrale in the Congo Basin or in the Okavango inland delta, the principle
30 threat is changing rainfall regimes due to climate variability and change (Weinzierl et al. 2016; Dargie
31 et al. 2017). Expansion of agriculture is not yet a major factor in these regions. In the Western
32 Amazon, extraction of non-timber forest products like the fruits of *Mauritia flexuosa* (morange palm)
33 and Suri worms are major sources of degradation that lead to losses of carbon stocks (Hergoualc'h et
34 al. 2017a).

35 The effects of peatland degradation on livelihoods have not been systematically characterised. In
36 places where plantation crops are driving the conversion of peat swamps, the financial benefits can be
37 considerable. One study in Indonesia found that the net present value of an oil palm plantation is
38 between USD 3,835 and 9,630 per ha to land owners (Butler et al. 2009). High financial returns are
39 creating the incentives for the expansion of smallholder production in peatlands. Smallholder
40 plantations extend over 22% of the peatlands in insular Southeast Asia compared to 27% for industrial
41 plantations (Miettinen et al. 2016). In places where income is generated from extraction of
42 marketable products, ecosystem degradation probably has a negative effect on livelihoods. For
43 example, the sale of fruits of *M. flexuosa* in some parts of the western Amazon constitutes as much as
44 80% of the winter income of many rural households, but information on trade values and value chains
45 of *M. flexuosa* is still sparse (Sousa et al. 2018; Virapongse et al. 2017).

46 There is little experience with peatland restoration in the tropics. Experience from northern latitudes
47 suggests that extensive damage and changes in hydrological conditions mean that restoration in many

1 cases is unachievable (Andersen et al. 2017). In the case of Southeast Asia, where peatlands form as
2 raised bogs, drainage leads to collapse of the dome and this collapse cannot be reversed by rewetting.
3 Nevertheless, efforts are underway to develop solutions or at least partial solutions in Southeast Asia,
4 for example, by the Indonesian Peatland Restoration Agency. The first step is to restore the
5 hydrological regime in drained peatlands and experiences with canal blocking and re-flooding of the
6 peat. These efforts have been only partially successful (Ritzema et al. 2014). Market incentives with
7 certification through the Roundtable on Sustainable Palm Oil have also not been particularly
8 successful as many concessions seek certification only after significant environmental degradation has
9 been accomplished (Carlson et al. 2017). Certification had no discernible effect on forest loss or fire
10 detection in peatlands in Indonesia. To date there is no documentation of restoration methods or
11 successes in many other parts of the tropics, but in situations where degradation does not involve
12 drainage, ecological restoration may be possible. In South America, for example, there is growing
13 interest in restoration of palm swamps, and as experiences are gained it will be important to document
14 success factors to inform successive efforts (Virapongse et al. 2017).

15 In higher latitudes where degraded peatlands have been drained, the most effective option to reduce
16 losses from these large organic carbon stocks is change hydrological conditions and increase soil
17 moisture and surface wetness (Regina et al. 2015). Long-term GHG monitoring in boreal sites has
18 demonstrated that rewetting and restoration noticeably reduce emissions compared to degraded
19 drained sites and can restore the carbon sink function when vegetation is re-established (Wilson et al.
20 2016; IPCC 2014a; Nugent et al. 2018) although restored ecosystems may not yet be as resilient as
21 their undisturbed counterparts (Wilson et al. 2016). Several studies have demonstrated the co-benefits
22 of rewetting specific degraded peatlands for biodiversity, carbon sequestration, (Parry et al. 2014;
23 Ramchunder et al. 2012; Renou-Wilson et al. 2018) and other ecosystem services such as
24 improvement of water storage and quality (Martin-Ortega et al. 2014) with beneficial consequences
25 for human well-being (Bonn et al. 2016; Parry et al. 2014).

26 **4.9.5 Biochar**

27 Biochar is organic matter that is carbonised by heating in an oxygen-limited environment, and used as
28 a soil amendment. The properties of biochar vary widely, dependent on the feedstock and the
29 conditions of production. Biochar could make a significant contribution to mitigating both land
30 degradation and climate change, simultaneously.

31 **4.9.5.1 Role of biochar in climate change mitigation**

32 Biochar is relatively resistant to decomposition compared with fresh organic matter or compost, so
33 represents a long-term C store (*very high confidence*). Biochars produced at higher temperature (>
34 450°C) and from woody material have greater stability than those produced at lower temperature
35 (300-450°C), and from manures (*very high confidence*) (Singh et al. 2012; Wang et al. 2016b).
36 Biochar stability is influenced by soil properties: biochar carbon can be further stabilised by
37 interaction with clay minerals and native soil organic matter (*medium evidence*) (Fang et al. 2015).
38 Biochar stability is estimated to range from decades to thousands of years, for different biochars in
39 different applications (Singh et al., 2015; Wang et al., 2016). Biochar stability decreases as ambient
40 temperature increases (*limited evidence*) (Fang et al. 2017).

41 Biochar can enhance soil carbon stocks through “negative priming”, in which rhizodeposits are
42 stabilised through sorption of labile C on biochar, and formation of biochar-organo-mineral
43 complexes (Weng et al. 2015, 2017, 2018; Wang et al. 2016b). Conversely, some studies show
44 increased turnover of native soil carbon (“positive priming”) due to enhanced soil microbial activity
45 induced by biochar. In clayey soils, positive priming is minor and short-lived compared to negative
46 priming effects, which dominate in the medium to long-term (Singh and Cowie 2014; Wang et al.
47 2016b). Negative priming has been observed particularly in loamy grassland soil (Ventura et al.

1 2015) and clay-dominated soils, whereas positive priming is reported in sandy soils (Wang et al.
2 2016b) and those with low C content (Ding et al. 2018).

3 Biochar can provide additional climate change mitigation by decreasing nitrous oxide (N₂O)
4 emissions from soil, due in part to decreased substrate availability for denitrifying organisms, related
5 to the molar H/C ratio of the biochar (Cayuela et al. 2015). However, this impact varies widely: meta-
6 analyses found an average decrease in N₂O emissions from soil of 30-54%, (Cayuela et al. 2015)
7 (Moore 2002; Borchard et al. 2019), although another study found no significant reduction in field
8 conditions when weighted by the inverse of the number of observations per site (Verhoeven et al.
9 2017). Biochar has been observed to reduce methane emissions from flooded soils, such as rice
10 paddies, though, as for N₂O, results vary between studies and increases have also been observed (He
11 et al. 2017; KAMMANN et al. 2017). Biochar has also been found to reduce methane uptake by
12 dryland soils, though the effect is small in absolute terms (Jeffery et al. 2016).

13 Additional climate benefits of biochar can arise through reduced N fertiliser requirements, due to
14 reduced losses of N through leaching and/or volatilization (Singh, Hatton, Balwant, & Cowie, 2010)
15 and enhanced biological nitrogen fixation (Van Zwieten et al. 2015); increased yields of crop, forage,
16 vegetable and tree species (Biederman and Stanley Harpole 2013), particularly in sandy soils and
17 acidic tropical soils (Simon et al. 2017) ; avoided GHG emissions from manure that would otherwise
18 be stockpiled, crop residues that would be burned or processing residues that would be landfilled; and
19 reduced GHG emissions from compost when biochar is added (Agyarko-Mintah et al. 2017; Wu et al.
20 2017a).

21 Climate benefits of biochar could be substantially reduced through reduction in albedo if biochar is
22 surface-applied at high rates to light-colored soils (Genesio et al. 2012; Bozzi et al. 2015; Woolf et al.
23 2010), or if black carbon dust is released (Genesio et al. 2016). Pelletizing or granulating biochar, and
24 applying below the soil surface or incorporating into the soil, minimises the release of black carbon
25 dust and reduces the effect on albedo (Woolf et al. 2010).

26 Biochar is a potential “negative emissions” technology: the thermochemical conversion of biomass to
27 biochar slows mineralisation of the biomass, delivering long term C storage; gases released during
28 pyrolysis can be combusted for heat or power, displacing fossil energy sources, and could be captured
29 and sequestered if linked with infrastructure for carbon capture and storage (Smith 2016). Studies of
30 the life cycle climate change impacts of biochar systems generally show emissions reduction in the
31 range 0.4 -1.2 t CO₂e t⁻¹ (dry) feedstock (Cowie et al. 2015). Use of biomass for biochar can deliver
32 greater benefits than use for bioenergy, if applied in a context where it delivers agronomic benefits
33 and/or reduces non-CO₂ GHG emissions (Ji et al. 2018; Woolf et al. 2010, 2018; Xu et al. 2019). A
34 global analysis of technical potential, in which biomass supply constraints were applied to protect
35 against food insecurity, loss of habitat and land degradation, estimated technical potential abatement
36 of 3.7 - 6.6 Gt CO₂e yr⁻¹ (including 2.6-4.6 GtCO₂e yr⁻¹ carbon stabilization), with theoretical
37 potential to reduce total emissions over the course of a century by 240 – 475 Gt CO₂e (Woolf et al.
38 2010). Fuss et al. 2018 propose a range of 0.5-2 GtCO₂e as the sustainable potential for negative
39 emissions through biochar. Mitigation potential of biochar is reviewed in Chapter 2.

40 **4.9.5.2 Role of biochar in management of land degradation**

41 Biochars generally have high porosity, high surface area and surface-active properties that lead to
42 high absorptive and adsorptive capacity, especially after interaction in soil (Joseph et al. 2010). As a
43 result of these properties, biochar could contribute to avoiding, reducing and reversing land
44 degradation through the following documented benefits:

- 45 • Improved nutrient use efficiency due to reduced leaching of nitrate and ammonium (e.g.
46 (Haider et al. 2017) and increased availability of phosphorus (P) in soils with high P fixation
47 capacity (Liu et al. 2018c), potentially reducing N and P fertiliser requirements.

- 1 • Management of heavy metals and organic pollutants: through reduced bioavailability of toxic
2 elements (O'Connor et al., 2018; Peng ; Deng, ; Peng, & Yue, 2018), by reducing availability,
3 through immobilization due to increased pH and redox effects (Rizwan et al. 2016) and
4 adsorption on biochar surfaces (Zhang et al. 2013) thus providing a means of remediating
5 contaminated soils, and enabling their utilisation for food production.
- 6 • Stimulation of beneficial soil organisms, including earthworms and mycorrhizal fungi (Thies
7 et al. 2015).
- 8 • Improved porosity and water holding capacity (Quin et al. 2014), particularly in sandy soils
9 (Omondi et al. 2016), enhancing microbial function during drought (Paetsch et al. 2018).
- 10 • Amelioration of soil acidification, through application of biochars with high pH and acid
11 neutralising capacity (Chan et al. 2008)(Van Zwieten et al. 2010).

12
13 Biochar systems can deliver a range of other co-benefits including destruction of pathogens and weed
14 propagules, avoidance of landfill, improved handling and transport of wastes such as sewage sludge,
15 management of biomass residues such as environmental weeds and urban greenwaste, reduction of
16 odors and management of nutrients from intensive livestock facilities, reduction in environmental N
17 pollution and protection of waterways. As a compost additive, biochar has been found to reduce
18 leaching and volatilisation of nutrients, increasing nutrient retention, through absorption and
19 adsorption processes (Joseph et al. 2018).

20 While many studies report positive responses, some studies have found negative or zero impacts on
21 soil properties or plant response (e.g. Kuppusamy, Thavamani, Megharaj, Venkateswarlu, & Naidu,
22 2016). The risk that biochar may enhance PAH in soil or sediments has been raised (Quilliam et al.
23 2013; Ojeda et al. 2016), but bioavailability of PAH in biochar has been shown to be very low (Hilber
24 et al. 2017) Pyrolysis of biomass leads to losses of volatile nutrients, especially N. While availability
25 of N and P in biochar is lower in biochar than in fresh biomass (Xu et al. 2016) the impact of biochar
26 on plant uptake is determined by the interactions between biochar, soil minerals and activity of
27 microorganisms (e.g. (Vanek and Lehmann 2015); (Nguyen et al. 2017). To avoid negative responses,
28 it is important to select biochar formulations to address known soil constraints, and to apply biochar
29 prior to planting (Nguyen et al., 2017). Nutrient enrichment improves the performance of biochar
30 from low nutrient feedstocks (Joseph et al. 2013). While there are many reports of biochar reducing
31 disease or pest incidence, there are also reports of nil or negative effects (Bonanomi et al. 2015).
32 Biochar may induce systemic disease resistance (e.g., Elad et al. 2011)), though (Viger et al. 2015)
33 reported down-regulation of plant defence genes, suggesting increased susceptibility to insect and
34 pathogen attack. Disease suppression where biochar is applied is associated with increased microbial
35 diversity and metabolic potential of the rhizosphere microbiome (Kolton et al. 2017). Differences in
36 properties related to feedstock (Bonanomi et al. 2018) and differential response to biochar dose, with
37 lower rates more effective (Frenkel et al. 2017) in contributing to variable disease responses.

38 Constraints to biochar adoption are high cost and limited availability due to limited large-scale
39 production; limited amount of unutilised biomass; and competition for land for growing biomass.
40 While early biochar research tended to use high rates of application (10 t ha⁻¹ or more) subsequent
41 studies have shown that biochar can be effective at lower rates especially when combined with
42 chemical or organic fertilisers (Joseph et al. 2013). Biochar can be produced at many scales and
43 levels of engineering sophistication, from simple cone kilns and cookstoves to large industrial scale
44 units processing several tonnes of biomass per hour (Lehmann and Stephen 2015). Substantial
45 technological development has occurred recently, though large-scale deployment is limited to date.

46 Governance of biochar is required to manage climate, human health and contamination risks
47 associated with biochar production in poorly-designed or operated facilities that release methane or
48 particulates (Downie et al. 2012)(Buss et al. 2015), to ensure quality control of biochar products, and

1 to ensure biomass is sourced sustainably and is uncontaminated. Measures could include labelling
2 standards, sustainability certification schemes and regulation of biochar production and use.
3 Governance mechanisms should be tailored to context, commensurate with risks of adverse outcomes.

4 In summary, application of biochar to soil can improve soil chemical, physical and biological
5 attributes, enhancing productivity and resilience to climate change, while also delivering climate
6 change mitigation through carbon sequestration and reduction in GHG emissions (*medium agreement,*
7 *robust evidence*). However, responses to biochar depend on biochar properties, in turn dependent on
8 feedstock and biochar production conditions, and the soil and crop to which it is applied. Negative or
9 nil results have been recorded. Agronomic and methane reduction benefits appear greatest in tropical
10 regions, where acidic soils predominate and suboptimal rates of lime and fertiliser are common, while
11 carbon stabilisation is greater in temperate regions. Biochar is most effective when applied in low
12 volumes to the most responsive soils and when properties are matched to the specific soil constraints
13 and plant needs. Biochar is thus a practice that has potential to address land degradation and climate
14 change simultaneously, while also supporting sustainable development. The potential of biochar is
15 limited by the availability of biomass for its production. Biochar production and use requires
16 regulation and standardisation to manage risks (*strong agreement*).

17 **4.9.6 Management of land degradation induced by tropical cyclones**

18 Tropical cyclones are normal disturbances that natural ecosystems have been affected by and
19 recovered from for millennia. Climate models mostly predict decreasing frequency of tropical
20 cyclones, but dramatically increasing intensity of the strongest storms as well as increasing rainfall
21 rates (Bacmeister et al. 2018; Walsh et al. 2016b). Large amplitude fluctuations in the frequency and
22 intensity complicate both the detection and attribution of tropical cyclones to climate change (Lin and
23 Emanuel 2016b). Yet, the intensity of high-intensity cyclones have increased and are expected to
24 increase further due to global climate change (Knutson et al. 2010; Bender et al. 2010; Vecchi et al.
25 2008; Bhatia et al. 2018; Tu et al. 2018; Sobel et al. 2016) (*medium agreement, robust evidence*).
26 Tropical cyclone paths are also shifting towards the poles increasing the area subject to tropical
27 cyclones (Sharmila and Walsh 2018; Lin and Emanuel 2016b). Climate change alone will affect the
28 hydrology of individual wetland ecosystems mostly through changes in precipitation and temperature
29 regimes with great global variability (Erwin 2009). Over the last seven decades, the speed at which
30 tropical cyclones move has decreased significantly as expected from theory, exacerbating the damage
31 on local communities from increasing rainfall amounts and high wind speed (Kossin 2018). Tropical
32 cyclones will accelerate changes in coastal forest structure and composition. The heterogeneity of
33 land degradation at coasts that are affected by tropical cyclones can be further enhanced by the
34 interaction of its components (for example, rainfall, wind speed, and direction) with topographic and
35 biological factors (for example, species susceptibility) (Luke et al. 2016).

36 Small Island Developing States (SIDS) are particularly affected by land degradation induced by
37 tropical cyclones, recent examples are Matthew (2016) in the Caribbean, and Pam (2015) and
38 Winston (2016) in the Pacific (Klöck and Nunn 2019; Handmer and Nalau 2019). Even if the Pacific
39 Ocean has experienced cyclones of unprecedented intensity in the recent years, their
40 geomorphological effects may not be unprecedented (Terry and Lau 2018).

41 Cyclone impacts on coastal areas is not restricted to SIDS, but a problem for all low-lying coastal
42 areas (Petzold and Magnan 2019). The Sundarban, one of the world's largest coastal wetlands, covers
43 about one million hectares between Bangladesh and India. Large areas of the Sundarban mangroves
44 have been converted into paddy fields over the past two centuries and more recently into shrimp farms
45 (Ghosh et al. 2015). In 2009 the cyclone Aila caused incremental stresses on the socioeconomic
46 conditions of the Sundarban coastal communities through rendering huge areas of land unproductive
47 for a long time (Abdullah et al. 2016). The impact of Aila was wide spread throughout the Sundarbans

1 mangroves showing changes between pre- and post-cyclonic period of 20-50% in the enhanced
2 vegetation index (Dutta et al. 2015). Although the magnitude of the effects of the Sundarban
3 mangroves derived from climate change is not yet defined (Payo et al. 2016; Loucks et al. 2010;
4 Gopal and Chauhan 2006; Ghosh et al. 2015; Chaudhuri et al. 2015). There is *high agreement* that the
5 joint effect of climate change and land degradation will be very negative for the area, strongly
6 affecting the environmental services provided by these forests, including the extinction of large
7 mammal species (Loucks et al. 2010). This changes in vegetation are mainly due to inundation and
8 erosion (Payo et al. 2016).

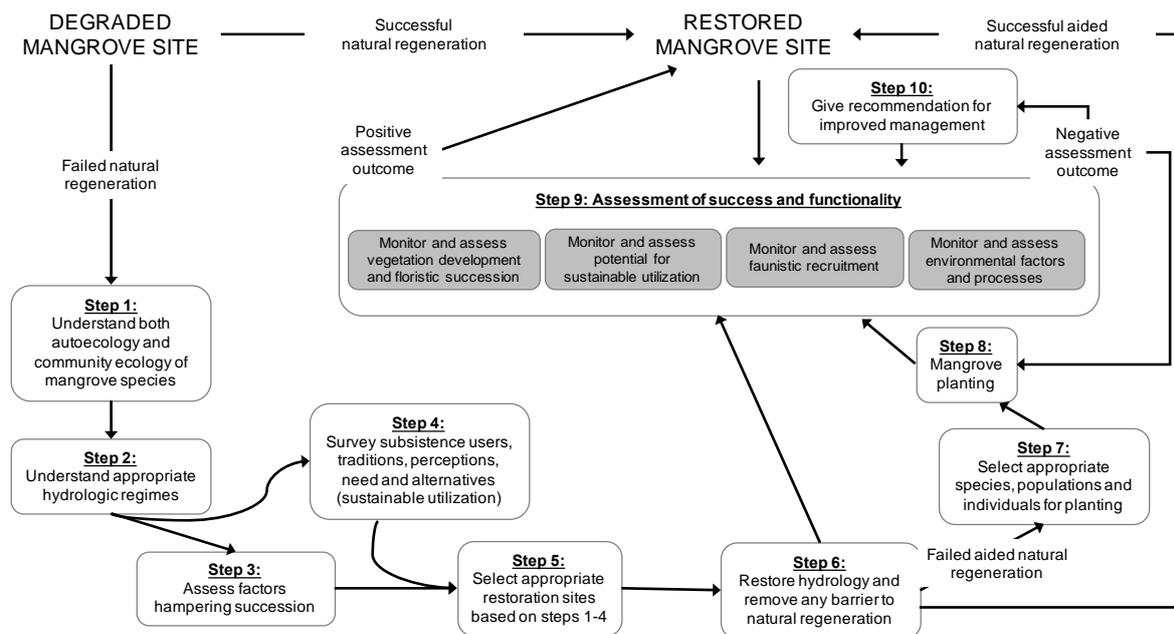
9 The tropical cyclone Nargis hit unexpectedly the Ayeyarwady River delta (Myanmar) in 2008 with
10 unprecedented and catastrophic damages to livelihoods, destruction of forests and erosion of fields
11 (Fritz et al. 2009) as well as eroding the shoreline 148 m compared with the long-term average (1974-
12 2015) of 0.62 m yr⁻¹. This is an example of the disastrous effects that changing cyclone paths can have
13 on areas previously not affected by cyclones (Fritz et al. 2010).

14 **4.9.6.1 Management of coastal wetlands**

15 Tropical cyclones mainly, but not exclusively, affect coastal regions, threatening maintenance of the
16 associated ecosystems, mangroves, wetlands, seagrasses, etc. This areas not only provide food, water
17 and shelter for fish, birds and other wildlife, but also provide important ecosystem services such as
18 water quality improvement, flood abatement and carbon sequestration (Meng et al. 2017).

19 Despite its importance coastal wetlands are listed amongst the most heavily damaged of natural
20 ecosystems worldwide. Starting in the 1990s, wetland restoration and re-creation became a “hotspot”
21 in the ecological research fields (Zedler 2000). The coastal wetland restoration and preservation is an
22 extremely cost-effective strategy for society, for example the preservation of coastal wetlands in the
23 USA provide storm protection services with the cost of 23.2 billion yr⁻¹ USD (Costanza et al. 2008).

24 There is a *high agreement* with *medium evidence* that the success of wetland restoration depends
25 mainly on the flow of the water through the system and the degree to which re-flooding occurs, the
26 disturbance regimes, and the control of invasive species (Burlakova et al. 2009; López-Rosas et al.
27 2013). The implementation of the Ecological Mangrove Rehabilitation (EMR) protocol (López-
28 Portillo et al. 2017) that includes monitoring and reporting tasks, has been proven to deliver
29 successful rehabilitation of wetland ecosystem services.



1
2 **Figure 4.10 Decision tree showing recommended steps and tasks to restore a mangrove wetland based on**
3 **original site conditions (Modified from Bosire et al. (2008))**

4 **4.9.7 Saltwater intrusion**

5 Current environmental changes, including climate change, have caused sea levels to rise worldwide,
6 particularly in tropical and subtropical regions (Fasullo and Nerem 2018). Combined with scarcity of
7 water in river channels, such rises have been instrumental in the intrusion of highly saline seawater
8 inland, posing a threat to coastal areas and an emerging challenge to land managers and policy
9 makers. Assessing the extent of salinisation due to sea water intrusion at a global scale nevertheless
10 remains challenging. Wicke et al. (2011) suggest that across the world, approximately 1.1 Gha of land
11 is affected by salt, with 14% of this categorised as forest, wetland or some other form of protected
12 area. Seawater intrusion is generally caused by: i) increased tidal activity, storm surges, cyclones and
13 sea storms due to changing climate, ii) heavy groundwater extraction or land use changes as a result
14 of changes in precipitation, and droughts/floods, iii) coastal erosion as a result of destruction of
15 mangrove forests and wetlands iv) construction of vast irrigation canals and drainage networks
16 leading to low river discharge in the deltaic region; and v) sea level rise contaminating nearby
17 freshwater aquifers as a result of subsurface intrusion (Uddameri et al. 2014).

18 The Indus delta, located in the south-eastern coast of Pakistan near Karachi in the North Arabian sea,
19 is one of the six largest estuaries in the world spanning an area of 600,000 ha. The Indus delta is a
20 clear example of seawater intrusion and land degradation due to local as well as up-country climatic
21 and environmental conditions (Rasul et al. 2012). Salinisation and waterlogging in the up-country
22 areas including provinces of Punjab and Sindh is, however, caused by the irrigation network and over-
23 irrigation (Qureshi 2011).

24 Such degradation takes the form of high soil salinity, inundation and waterlogging, erosion and
25 freshwater contamination. The inter-annual variability of precipitation with flooding conditions in
26 some years and drought conditions in others has caused variable river flows and sediment runoff
27 below Kotri barrage (about 200 km upstream of the Indus delta). This has affected hydrological
28 processes in the lower reaches of the river and the delta, contributing to the degradation (Rasul et al.
29 2012).

30 Over 480,000 ha of fertile land is now affected by sea water intrusion, wherein eight coastal
31 subdivisions of the districts of Badin and Thatta are mostly affected (Chandio et al. 2011). A very
32 high intrusion rate of $0.179 \pm 0.0315 \text{ km yr}^{-1}$, based on the analysis of satellite data, was observed in

1 the Indus delta during the past 10 years (2004–2015) (Kalhor et al. 2016). The area of agricultural
2 crops under cultivation has been declining with economic losses of millions of USD (IUCN 2003).
3 Crop yields have reduced due to soil salinity, in some places failing entirely. Soil salinity varies
4 seasonally, depending largely on the river discharge: during the wet season (August 2014), salinity
5 (0.18 mg L^{-1}) reached 24 km upstream while during the dry season (May 2013), it reached 84 km
6 upstream (Kalhor et al. 2016). The freshwater aquifers have also been contaminated with sea water
7 rendering them unfit for drinking or irrigation purposes. Lack of clean drinking water and sanitation
8 causes widespread diseases, of which diarrhoea is most common (IUCN 2003).

9 Lake Urmia in northwest Iran, the second largest saltwater lake in the world and the habitat for
10 endemic Iranian brine shrimp, *Artemia urmiana*, has also been affected by salty water intrusion.
11 During a 17-year period between 1998 and 2014, human disruption including agriculture and years of
12 dam building affected the natural flow of freshwater as well as salty sea water in the surrounding area
13 of Lake Urmia. Water quality has also been adversely affected, with salinity fluctuating over time, but
14 in recent years reaching a maximum of 340 g L^{-1} (similar to levels in the Dead Sea). This has rendered
15 the underground water unfit for drinking and agricultural purposes and risky to human health and
16 livelihoods. Adverse impacts of global climate change as well as direct human impacts have caused
17 changes in land use, overuse of underground water resources and construction of dams over rivers
18 which resulted in the drying-up of the lake in large part. This condition created sand, dust and salt
19 storms in the region which affected many sectors including agriculture, water resources, rangelands,
20 forests and health, and generally presented desertification conditions around the lake (Karbassi et al.
21 2010; Marjani and Jamali 2014; Shadkam et al. 2016).

22 Rapid irrigation expansion in the basin has, however, indirectly contributed to inflow reduction.
23 Annual inflow to Lake Urmia has dropped by 48% in recent years. About three fifths of this change
24 was caused by climate change and two fifths by water resource development and agriculture (Karbassi
25 et al. 2010; Marjani and Jamali 2014; Shadkam et al. 2016).

26 In the drylands of Mexico, intensive production of irrigated wheat and cotton using groundwater
27 (Halvorson et al. 2003) resulted in sea water intrusion into the aquifers of La Costa de Hermosillo, a
28 coastal agricultural valley at the center of Sonora Desert in Northwestern Mexico. Production of these
29 crops in 1954 was on 64,000 ha of cultivated area, increasing to 132,516 ha in 1970, but decreasing to
30 66,044 ha in 2009 as a result of saline intrusion from the Gulf of California (Romo-Leon et al. 2014).
31 In 2003, only 15% of the cultivated area was under production, with around 80,000 ha abandoned due
32 to soil salinisation whereas in 2009, around 40,000 ha was abandoned (Halvorson et al. 2003; Romo-
33 Leon et al. 2014). Salinisation of agricultural soils could be exacerbated by climate change, as
34 Northwestern Mexico is projected to be warmer and drier under climate change scenarios (IPCC
35 2013a).

36 In other countries, intrusion of seawater is exacerbated by destruction of mangrove forests.
37 Mangroves are important coastal ecosystems that provide spawning bed for fish, timber for building,
38 livelihoods to dependent communities, act as barriers against coastal erosion, storm surges, tropical
39 cyclones and tsunamis (Kalhor et al. 2017) and are among the most carbon-rich stocks on Earth
40 (Atwood et al. 2017). They nevertheless face a variety of threats: climatic (storm surges, tidal
41 activities, high temperatures) and human (coastal developments, pollution, deforestation, conversion
42 to aquaculture, rice culture, oil palm plantation), leading to declines in their areas. In Pakistan, using
43 remote sensing (RS), the mangrove forest cover in the Indus delta decreased from 260,000 ha in
44 1980s to 160,000 ha in 1990 (Chandio et al. 2011). Based on remotely sensed data, a sharp decline in
45 the mangrove area was also found in the arid coastal region of Hormozgan province in southern Iran
46 during 1972, 1987 and 1997 (Etemadi et al. 2016). Myanmar has the highest rate (about $1\% \text{ yr}^{-1}$) of
47 mangrove deforestation in the world (Atwood et al. 2017). Regarding global loss of carbon stored in
48 the mangrove due to deforestation, four countries exhibited high levels of loss: Indonesia (3,410 Gg

1 CO₂ yr⁻¹), Malaysia (1,288 GgCO₂ yr⁻¹), US (206 Gg CO₂ yr⁻¹) and Brazil (186 GgCO₂ yr⁻¹). Only in
2 Bangladesh and Guinea Bissau there was no decline in the mangrove area from 2000 to 2012
3 (Atwood et al. 2017).

4 Frequency and intensity of average tropical cyclones will continue to increase (Knutson et al. 2015)
5 and global sea level will continue to rise. The IPCC (2013) projected with *medium confidence* that sea
6 level in the Asia Pacific region will rise from 0.4 to 0.6 m, depending on the emission pathway, by the
7 end of this century. Adaptation measures are urgently required to protect the world's coastal areas
8 from further degradation due to saline intrusion. A viable policy framework is needed to ensure the
9 environmental flows to deltas in order to repulse the intruding seawater.

10 **4.9.8 Avoiding coastal maladaptation**

11 Coastal degradation—for example, beach erosion, coastal squeeze, and coastal biodiversity loss—as a
12 result of rising sea levels is a major concern for low lying coasts and small islands (*high confidence*).
13 The contribution of climate change to increased coastal degradation has been well documented in
14 AR5 (Nurse et al. 2014; Wong et al. 2014) and is further discussed in Section 4.4.1.3. as well as in the
15 IPCC Special Report on the Ocean and Cryosphere in a Changing Climate (SROCC). However,
16 coastal degradation can also be indirectly induced by climate change as the result of adaptation
17 measures that involve changes to the coastal environment, for example, coastal protection measures
18 against increased flooding and erosion due to sea level rise and storm surges transforming the natural
19 coast to a ‘stabilised’ coastline (Cooper and Pile 2014; French 2001). Every kind of adaptation
20 response option is context-dependent, and, in fact, sea walls play an important role for adaptation in
21 many places. Nonetheless, there are observed cases where the construction of sea walls can be
22 considered ‘maladaptation’ (Barnett and O’Neill 2010; Magnan et al. 2016) by leading to increased
23 coastal degradation, such as in the case of small islands, where due to limitations of space coastal
24 retreat is less of an option than in continental coastal zones. There is emerging literature on the
25 implementation of alternative coastal protection measures and mechanisms on small islands to avoid
26 coastal degradation induced by sea walls (e.g., Mycoo and Chadwick 2012; Sovacool 2012).

27 In many cases, increased rates of coastal erosion due to the construction of sea walls are the result of
28 the negligence of local coastal morphological dynamics and natural variability as well as the interplay
29 of environmental and anthropogenic drivers of coastal change (*medium evidence, high agreement*).
30 Sea walls in response to coastal erosion may be ill-suited for extreme wave heights under cyclone
31 impacts and can lead to coastal degradation by keeping overflowing sea water from flowing back into
32 the sea, and therefore affect the coastal vegetation through saltwater intrusion, as observed in Tuvalu
33 (Government of Tuvalu 2006; Wairiu 2017). Similarly, in Kiribati, poor construction of sea walls has
34 resulted in increased erosion and inundation of reclaimed land (Donner 2012; Donner and Webber
35 2014). In the Comoros and Tuvalu, sea walls have been constructed from climate change adaptation
36 funds and ‘often by international development organizations seeking to leave tangible evidence of
37 their investments’ (Marino and Lazrus 2015, p. 344). In these cases, they have even increased coastal
38 erosion, due to poor planning and the negligence of other causes of coastal degradation, such as sand
39 mining (Marino and Lazrus 2015; Betzold and Mohamed 2017; Ratter et al. 2016). On the Bahamas,
40 the installation of sea walls as a response to coastal erosion in areas with high wave action has led to
41 the contrary effect and even increased sand loss in those areas (Sealey 2006). The reduction of natural
42 buffer zones—i.e., beaches and dunes—due to vertical structures, such as sea walls, increased the
43 impacts of tropical cyclones on Reunion Island (Duvat et al. 2016). Such a process of ‘coastal
44 squeeze’ (Pontee 2013) also results in the reduction of intertidal habitat zones, such as wetlands and
45 marshes (Linham and Nicholls 2010). Coastal degradation resulting from the construction of sea
46 walls, however, is not only observed in Small Island Developing States (SIDS), as described above,
47 but also on islands in the Global North, for example, the North Atlantic (Muir et al. 2014; Young et al.
48 2014; Cooper and Pile 2014; Bush 2004).

1 The adverse effects of coastal protection measures may be avoided by the consideration of local
2 social-ecological dynamics, including the critical studying of diverse drivers of ongoing shoreline
3 changes, and the according implementation of locally adequate coastal protection options (French
4 2001; Duvat 2013). Critical elements for avoiding maladaptation include profound knowledge of local
5 tidal regimes, availability of relative sea level rise scenarios and projections for extreme water levels.
6 Moreover, the downdrift effects of sea walls need to be considered, since undefended coasts may be
7 exposed to increased erosion (Linham and Nicholls 2010). In some cases, it may be possible to keep
8 intact and restore natural buffer zones as an alternative to the construction of hard engineering
9 solutions. Otherwise, changes in land-use, building codes, or even coastal realignment can be an
10 option in order to protect and avoid the loss of the buffer function of beaches (Duvat et al. 2016;
11 Cooper and Pile 2014). Examples of Barbados show that combinations of hard and soft coastal
12 protection approaches can be sustainable and reduce the risk of coastal ecosystem degradation while
13 keeping the desired level of protection for coastal users (Mycoo and Chadwick 2012). Nature-based
14 solutions and approaches such as ‘building with nature’ (Slobbe et al. 2013) may allow for more
15 sustainable coastal protection mechanisms and avoid coastal degradation. Examples from the
16 Maldives, several Pacific islands and the North Atlantic show the importance of the involvement of
17 local communities in coastal adaptation projects, considering local skills, capacities, as well as
18 demographic and socio-political dynamics, in order to ensure the proper monitoring and maintenance
19 of coastal adaptation measures (Sovacool 2012; Muir et al. 2014; Young et al. 2014; Buggy and
20 McNamara 2016; Petzold 2016).

21 **4.10 Knowledge gaps and key uncertainties**

22 The co-benefits of improved land management, such as mitigation of climate change, increased
23 climate resilience of agriculture, and impacts on rural areas/societies are well-known in theory but
24 there is a lack of a coherent and systematic global inventory of such integrated efforts. Both successes
25 and failures are important to document systematically.

26 Efforts to reduce climate change through land-demanding mitigation actions aimed at removing
27 atmospheric carbon, such as afforestation, reforestation, bioenergy crops, intensification of land
28 management and plantation forestry can adversely affect land conditions and lead to degradation.
29 However, they may also lead to avoidance, reduction and reversal of degradation. Regionally
30 differentiated, socially and ecologically appropriate sustainable land management strategies need to
31 be identified, implemented, monitored and the results communicated widely to ensure climate
32 effective outcomes.

33 Impacts of new technologies on land degradation and their social and economic ramifications need
34 more research.

35 Improved quantification of the global extent, severity and rates of land degradation by combining
36 remote sensing with a systematic use of ancillary data is a priority. The current attempts need a better
37 scientific underpinning and appropriate funding.

38 Land degradation is defined using multiple criteria but the definition does not provide thresholds or
39 the magnitude of acceptable change. In practice, human interactions with land will result in a variety
40 of changes, some may contribute positively to one criterion while adversely affecting another.
41 Research is required on the magnitude of impacts and the resulting trade-offs. Given the urgent need
42 to remove carbon from the atmosphere and to reduce climate change impacts, it is important to reach
43 agreement on what level of reduction in one criterion (biological productivity, ecological integrity)
44 may be acceptable for a given increase in another criterion (ecological integrity, biological
45 productivity)?

1 Attribution of land degradation to the underlying drivers is a challenge because a complex web of
2 causality rather than simple cause-effect relationships. Also, diverging views on land degradation in
3 relation to other challenges is hampering such efforts.

4 A more systematic treatment of the views and experiences of land users would be useful in land
5 degradation studies.

6 Much research has tried to understand how social and ecological systems are affected by a particular
7 stressor, for example drought, heat, or waterlogging. But less research has tried to understand how
8 such systems are affected by several simultaneous stressors – which of course is more realistic in the
9 context of climate change (Mittler 2006).

10 More realistic modelling of carbon dynamics, including better appreciation of belowground biota,
11 would help us to better quantify the role of soils and soil management for soil carbon sequestration.

12

13 **Frequently Asked Questions**

14 **FAQ 4.1 How do climate change and land degradation interact with land use?**

15 Climate change, land degradation, and land use are linked in a complex web of causality. One
16 important impact of climate change (e.g. flood and drought) on land degradation is that increasing
17 global temperatures intensify the hydrological cycle resulting in more intense rainfall, which is an
18 important driver of soil erosion. This means that sustainable land management (SLM) becomes even
19 more important with climate change. Land-use change in the form of clearing of forest for rangeland
20 and cropland (e.g., for provision of bio-fuels), and cultivation of peat soils, is a major source of
21 greenhouse gas emission from both biomass and soils. Many SLM practices (e.g., agroforestry,
22 shifting perennial crops, restoration, etc.) increase carbon content of soil and vegetation cover and
23 hence provide both local and immediate adaptation benefits combined with global mitigation benefits
24 in the long term, while providing many social and economic co-benefits. Avoiding, reducing and
25 reversing land degradation has a large potential to mitigate climate change and help communities to
26 adapt to climate change.

27

28 **FAQ 4.2 How does climate change affect land-related ecosystem services and 29 biodiversity?**

30 Climate change will affect land-related ecosystem services (e.g. pollination, resilience to extreme
31 climate events, water yield, soil conservation, carbon storage, etc.) and biodiversity, both directly and
32 indirectly. The direct impacts range from subtle reductions or enhancements of specific services, such
33 as biological productivity, resulting from changes in temperature, temperature variability or rainfall,
34 to complete disruption and elimination of services. Disruptions of ecosystem services can occur where
35 climate change causes transitions from one biome to another, e.g., forest to grassland as a result of
36 changes in water balance or natural disturbance regimes. Climate change will result in range shifts
37 and, in some cases, extinction of species. Climate change can also alter the mix of land-related
38 ecosystem services, such as groundwater recharge, purification of water, and flood protection. While
39 the net impacts are specific to ecosystem types, ecosystem services and time, there is an asymmetry of
40 risk such that overall impacts of climate change are expected to reduce ecosystem services. Indirect
41 impacts of climate change on land-related ecosystem services include those that result from changes
42 in human behavior, including potential large-scale human migrations or the implementation of
43 afforestation, reforestation or other changes in land management, which can have positive or negative
44 outcomes on ecosystem services.

45

1 References

- 2 Abatzoglou, J. T., and A. P. Williams, 2016: Impact of anthropogenic climate change on wildfire
3 across western US forests. *Proc. Natl. Acad. Sci. U. S. A.*, **113**, 11770–11775,
4 doi:10.1073/pnas.1607171113. <http://www.ncbi.nlm.nih.gov/pubmed/27791053> (Accessed
5 March 16, 2019).
- 6 Abbott, B. W., and Coauthors, 2016: Biomass offsets little or none of permafrost carbon release from
7 soils, streams, and wildfire: an expert assessment. *Environ. Res. Lett.*, **11**, 034014,
8 doi:10.1088/1748-9326/11/3/034014. [http://stacks.iop.org/1748-
9 9326/11/i=3/a=034014?key=crossref.66df61179f78005f9999c2637a68bf9e](http://stacks.iop.org/1748-9326/11/i=3/a=034014?key=crossref.66df61179f78005f9999c2637a68bf9e) (Accessed March
10 18, 2019).
- 11 Abdi, A. M., N. Boke-Olén, H. Jin, L. Eklundh, T. Tagesson, V. Lehsten, and J. Ardö, 2019: First
12 assessment of the plant phenology index (PPI) for estimating gross primary productivity in
13 African semi-arid ecosystems. *Int. J. Appl. Earth Obs. Geoinf.*, **78**, 249–260,
14 doi:10.1016/J.JAG.2019.01.018.
15 <https://www.sciencedirect.com/science/article/pii/S0303243418311085> (Accessed March 22,
16 2019).
- 17 Abdulai, I., P. Vaast, M. P. Hoffmann, R. Asare, L. Jassogne, P. Van Asten, R. P. Rötter, and S.
18 Graefe, 2018: Cocoa agroforestry is less resilient to sub-optimal and extreme climate than cocoa
19 in full sun. *Glob. Chang. Biol.*, **24**, 273–286, doi:10.1111/gcb.13885.
20 <http://doi.wiley.com/10.1111/gcb.13885> (Accessed November 1, 2018).
- 21 Abdullah, A. N. M., K. K. Zander, B. Myers, N. Stacey, and S. T. Garnett, 2016: A short-term
22 decrease in household income inequality in the Sundarbans, Bangladesh, following Cyclone
23 Aila. *Nat. Hazards*, **83**, 1103–1123, doi:10.1007/s11069-016-2358-1.
24 <http://link.springer.com/10.1007/s11069-016-2358-1> (Accessed June 19, 2018).
- 25 Abraha, M., S. K. Hamilton, J. Chen, and G. P. Robertson, 2018: Ecosystem carbon exchange on
26 conversion of Conservation Reserve Program grasslands to annual and perennial cropping
27 systems. *Agric. For. Meteorol.*, **253–254**, 151–160, doi:10.1016/J.AGRFORMET.2018.02.016.
28 <https://www.sciencedirect.com/science/article/pii/S0168192318300662> (Accessed May 17,
29 2018).
- 30 Abu, K., and Coauthors, 2017: How temporal patterns in rainfall determine the geomorphology and
31 carbon fluxes of tropical peatlands. *Proc. Natl. Acad. Sci.*, 201701090,
32 doi:10.1073/pnas.1701090114.
- 33 Achat, D. L., M. Fortin, G. Landmann, B. Ringeval, and L. Augusto, 2015: Forest soil carbon is
34 threatened by intensive biomass harvesting. *Sci. Rep.*, **5**, 15991, doi:10.1038/srep15991.
35 <http://www.nature.com/articles/srep15991> (Accessed February 27, 2019).
- 36 Adefuye, B. O., O. Odusan, T. I. Runsewe-Abiodun, T. Olowonyo, B. Bodunde, K. Alabi, and P. O.
37 Adefuye, 2007: Practice and perception of biomass fuel use and its health effects among
38 residents in a sub urban area of southern Nigeria. a qualitative study. *Niger. Hosp. Pract.*, **22**,
39 48–54. <https://www.ajol.info/index.php/nhp/article/view/180452> (Accessed April 10, 2019).
- 40 Adeogun, A. G., B. A. Ibitoye, A. W. Salami, and G. T. Ighag, 2018: Sustainable management of
41 erosion prone areas of upper watershed of Kainji hydropower dam, Nigeria. *J. King Saud Univ. -
42 Eng. Sci.*, doi:10.1016/J.JKSUES.2018.05.001.
43 <https://www.sciencedirect.com/science/article/pii/S1018363918300321> (Accessed March 8,
44 2019).
- 45 Adger, N. W., and Coauthors, 2014: Human Security. *Climate Change 2014 Impacts, Adaptation, and
46 Vulnerability*, C.B. Field and V.R. Barros, Eds., Cambridge University Press, Cambridge, UK
47 and New York, USA, 755–791
48 [https://s3.amazonaws.com/academia.edu.documents/36898556/WGIIAR5-
49 Chap12_FINAL.pdf?AWSAccessKeyId=AKIAIWOWYYGZ2Y53UL3A&Expires=151489659](https://s3.amazonaws.com/academia.edu.documents/36898556/WGIIAR5-Chap12_FINAL.pdf?AWSAccessKeyId=AKIAIWOWYYGZ2Y53UL3A&Expires=151489659)

- 1 7&Signature=IBxRb2IGBKXQluH5y%2F%2BNuOIJJwo%3D&response-content-
2 disposition=inline%3B filename%3DHuman_Security_WGII_Cha (Accessed January 2, 2018).
- 3 Adger, W. N., and Coauthors, 2009: Are there social limits to adaptation to climate change? *Clim.*
4 *Change*, **93**, 335–354, doi:10.1007/s10584-008-9520-z.
5 <http://link.springer.com/10.1007/s10584-008-9520-z> (Accessed November 2, 2018).
- 6 Adimassu, Z., S. Langan, and R. Johnston, 2016: Understanding determinants of farmers' investments
7 in sustainable land management practices in Ethiopia: review and synthesis. *Environ. Dev.*
8 *Sustain.*, **18**, 1005–1023, doi:10.1007/s10668-015-9683-5.
9 <http://link.springer.com/10.1007/s10668-015-9683-5> (Accessed April 17, 2019).
- 10 Agarwal, B., 1997: Environmental Action, Gender Equity and Women's Participation. *Dev. Change*,
11 **28**, 1–44, doi:10.1111/1467-7660.00033. <http://doi.wiley.com/10.1111/1467-7660.00033>
12 (Accessed April 16, 2018).
- 13 Agyarko-Mintah, E., A. Cowie, B. P. Singh, S. Joseph, L. Van Zwieten, A. Cowie, S. Harden, and R.
14 Smillie, 2017: Biochar increases nitrogen retention and lowers greenhouse gas emissions when
15 added to composting poultry litter. *Waste Manag.*, **61**, 138–149,
16 doi:10.1016/j.wasman.2016.11.027.
17 <https://linkinghub.elsevier.com/retrieve/pii/S0956053X16306869>.
- 18 Aleman, J. C., M. A. Jarzyna, and A. C. Staver, 2018: Forest extent and deforestation in tropical
19 Africa since. *Nat. Ecol. Evol.*, **2**, 26–33, doi:10.1038/s41559-017-0406-1.
- 20 Alkemade, R., M. van Oorschot, L. Miles, C. Nellemann, M. Bakkenes, and B. ten Brink, 2009:
21 GLOBIO3: A Framework to Investigate Options for Reducing Global Terrestrial Biodiversity
22 Loss. *Ecosystems*, **12**, 374–390, doi:10.1007/s10021-009-9229-5.
23 <http://link.springer.com/10.1007/s10021-009-9229-5> (Accessed March 18, 2019).
- 24 Allen, C. D., and Coauthors, 2002: Ecological Restoration of Southwestern Ponderosa Pine
25 Ecosystems: A Broad Perspective. *Ecol. Appl.*, **12**, 1418–1433, doi:10.2307/3099981.
- 26 —, and Coauthors, 2010: A global overview of drought and heat-induced tree mortality reveals
27 emerging climate change risks for forests. *For. Ecol. Manage.*, **259**, 660–684,
28 doi:10.1016/J.FORECO.2009.09.001.
29 <https://www.sciencedirect.com/science/article/pii/S037811270900615X> (Accessed March 5,
30 2018).
- 31 Allen, D. E., B. P. Singh, and R. C. Dalal, 2011: Soil Health Indicators Under Climate Change: A
32 Review of Current Knowledge. Springer, Berlin, Heidelberg, 25–45
33 http://link.springer.com/10.1007/978-3-642-20256-8_2 (Accessed November 5, 2017).
- 34 Allen, H. M., J. K. Pumpa, and G. D. Batten, 2001: Effect of frost on the quality of samples of Janz
35 wheat. *Aust. J. Exp. Agric.*, **41**, 641, doi:10.1071/EA00187.
36 [http://www.publish.csiro.au/journals/abstractHTML.cfm?J=EA&V=41&I=5&F=EA00187abs.X](http://www.publish.csiro.au/journals/abstractHTML.cfm?J=EA&V=41&I=5&F=EA00187abs.XML)
37 [ML](http://www.publish.csiro.au/journals/abstractHTML.cfm?J=EA&V=41&I=5&F=EA00187abs.XML) (Accessed May 31, 2018).
- 38 Allison, M., and Coauthors, 2016: Global Risks and Research Priorities for Coastal Subsidence. *Eos*
39 *(Washington, DC)*, **97**, doi:10.1029/2016EO055013. [https://eos.org/features/global-risks-and-](https://eos.org/features/global-risks-and-research-priorities-for-coastal-subsidence)
40 [research-priorities-for-coastal-subsidence](https://eos.org/features/global-risks-and-research-priorities-for-coastal-subsidence) (Accessed April 7, 2019).
- 41 Almagro, A., P. T. S. Oliveira, M. A. Nearing, and S. Hagemann, 2017: Projected climate change
42 impacts in rainfall erosivity over Brazil. *Sci. Rep.*, **7**, 8130, doi:10.1038/s41598-017-08298-y.
43 <http://www.nature.com/articles/s41598-017-08298-y> (Accessed December 5, 2017).
- 44 Alongi, D. M., 2015: The Impact of Climate Change on Mangrove Forests. *Curr. Clim. Chang.*
45 *Reports*, **1**, 30–39, doi:10.1007/s40641-015-0002-x. [http://link.springer.com/10.1007/s40641-](http://link.springer.com/10.1007/s40641-015-0002-x)
46 [015-0002-x](http://link.springer.com/10.1007/s40641-015-0002-x) (Accessed March 16, 2019).
- 47 Altieri, M. A., and P. Koohafkan, 2008: *Enduring Farms: Climate Change, Smallholders and*
48 *Traditional Farming Communities*. Penang, Malaysia, 72 pp.

- 1 http://sa.indiaenvironmentportal.org.in/files/Enduring_Farms.pdf.
- 2 —, and C. I. Nicholls, 2017: The adaptation and mitigation potential of traditional agriculture in a
3 changing climate. *Clim. Change*, **140**, 33–45, doi:10.1007/s10584-013-0909-y.
4 <http://link.springer.com/10.1007/s10584-013-0909-y> (Accessed October 20, 2017).
- 5 Altor, A. E., and W. J. Mitsch, 2006: Methane flux from created riparian marshes: Relationship to
6 intermittent versus continuous inundation and emergent macrophytes. *Ecol. Eng.*, **28**, 224–234,
7 doi:10.1016/j.ecoleng.2006.06.006.
- 8 Amanambu, A. C., L. Li, C. N. Egbinola, O. A. Obarein, C. Mupenzi, and D. Chen, 2019: Spatio-
9 temporal variation in rainfall-runoff erosivity due to climate change in the Lower Niger Basin,
10 West Africa. *CATENA*, **172**, 324–334, doi:10.1016/J.CATENA.2018.09.003.
11 <https://www.sciencedirect.com/science/article/pii/S0341816218303709> (Accessed March 13,
12 2019).
- 13 Amichev, B. Y., W. A. Kurz, C. Smyth, and K. C. J. Van Rees, 2012: The carbon implications of
14 large-scale afforestation of agriculturally marginal land with short-rotation willow in
15 Saskatchewan. *GCB Bioenergy*, **4**, doi:10.1111/j.1757-1707.2011.01110.x.
- 16 Amugune, I., P. Cerutti, H. Baral, S. Leonard, and C. Martius, 2017: *Small flame but no fire: Wood*
17 *fuel in the (Intended) Nationally Determined ... - Amugune, I., Cerutti, P.O., Baral, H., Leonard,*
18 *S., Martius, C. - Google Books.* Bogor, Indonesia, 35 pp.
19 [https://books.google.se/books?hl=en&lr=&id=CRZQDwAAQBAJ&oi=fnd&pg=PR3&dq=Amu
20 gune+I,+Cerutti+P,+Baral+H,+Leonard+S+and+Martius+C&ots=plDorqu9zX&sig=pVljVjzzw
21 8aj8wmOqCbzhHNpEHE&redir_esc=y#v=onepage&q=Amugune I%2C Cerutti P%2C Baral
22 H%2C Leonard](https://books.google.se/books?hl=en&lr=&id=CRZQDwAAQBAJ&oi=fnd&pg=PR3&dq=Amugune+I,+Cerutti+P,+Baral+H,+Leonard+S+and+Martius+C&ots=plDorqu9zX&sig=pVljVjzzw8aj8wmOqCbzhHNpEHE&redir_esc=y#v=onepage&q=Amugune%20Cerutti%20Baral%20Leonard) (Accessed November 2, 2018).
- 23 Andela, N., Y. Y. Liu, A. I. J. M. Van Dijk, R. A. M. De Jeu, T. R. McVicar, A. I. J. M. van Dijk, R.
24 A. M. De Jeu, and T. R. McVicar, 2013: *Global changes in dryland vegetation dynamics (1988-*
25 *2008) assessed by satellite remote sensing: Comparing a new passive microwave vegetation*
26 *density record with reflective greenness data.* <https://www.biogeosciences.net/10/6657/2013/>
27 (Accessed September 26, 2018).
- 28 Anderegg, W. R. L., J. A. Berry, and C. B. Field, 2012: Linking definitions, mechanisms, and
29 modeling of drought-induced tree death. *Trends Plant Sci.*, **17**, 693–700,
30 doi:10.1016/J.TPLANTS.2012.09.006.
31 <https://www.sciencedirect.com/science/article/pii/S1360138512002130#bib0010> (Accessed
32 March 5, 2018).
- 33 Andersen, R., C. Farrell, M. Graf, F. Muller, E. Calvar, P. Frankard, S. Caporn, and P. Anderson,
34 2017: An overview of the progress and challenges of peatland restoration in Western Europe.
35 *Restor. Ecol.*, **25**, 271–282, doi:10.1111/rec.12415.
- 36 Anderson, R. G., and Coauthors, 2011: Biophysical considerations in forestry for climate protection.
37 *Front. Ecol. Environ.*, **9**, 174–182, doi:10.1890/090179. <http://doi.wiley.com/10.1890/090179>
38 (Accessed December 20, 2017).
- 39 Anderson, R. L., D. R. Foster, and G. Motzkin, 2003: Integrating lateral expansion into models of
40 peatland development in temperate New England. *J. Ecol.*, **91**, 68–76, doi:10.1046/j.1365-
41 2745.2003.00740.x. <http://doi.wiley.com/10.1046/j.1365-2745.2003.00740.x> (Accessed
42 December 27, 2017).
- 43 Andersson, E., S. Brogaard, and L. Olsson, 2011: The Political Ecology of Land Degradation. *Annu.*
44 *Rev. Environ. Resour.*, **36**, 295–319, doi:10.1146/annurev-environ-033110-092827.
45 <http://www.annualreviews.org/doi/10.1146/annurev-environ-033110-092827> (Accessed April
46 16, 2018).
- 47 Angelsen, A., P. Jagger, R. Babigumira, N. J. Hogarth, S. Bauch, J. Börner, and C. Smith-Hall, 2014:
48 Environmental Income and Rural Livelihoods: A Global-Comparative Analysis. *World Dev.*, **64**,
49 S12–S28, doi:10.1016/J.WORLDDEV.2014.03.006.

- 1 <https://www.sciencedirect.com/science/article/pii/S0305750X14000722?via%3Dihub> (Accessed
2 March 8, 2019).
- 3 Angelsen, A., C. Martius, V. de Sy, A. E. Duchelle, A. M. Larson, and T. T. Pham, 2018:
4 Transforming REDD+: Lessons and new directions. Center for International Forestry Research
5 (CIFOR) <https://www.cifor.org/online-library/browse/view-publication/publication/7045.html>
6 (Accessed March 26, 2019).
- 7 Antwi-Agyei, P., A. J. Dougill, and L. C. Stringer, 2015: Impacts of land tenure arrangements on the
8 adaptive capacity of marginalized groups: The case of Ghana's Ejura Sekyedumase and Bongo
9 districts. *Land use policy*, **49**, 203–212, doi:10.1016/J.LANDUSEPOL.2015.08.007.
10 <https://www.sciencedirect.com/science/article/pii/S0264837715002422> (Accessed May 16,
11 2018).
- 12 Aragüés, R., E. T. Medina, W. Zribi, I. Clavería, J. Álvaro-Fuentes, and J. Faci, 2015: Soil
13 salinization as a threat to the sustainability of deficit irrigation under present and expected
14 climate change scenarios. *Irrig. Sci.*, **33**, 67–79, doi:10.1007/s00271-014-0449-x.
15 <http://link.springer.com/10.1007/s00271-014-0449-x> (Accessed March 18, 2019).
- 16 Arnáez, J., N. Lana-Renault, T. Lasanta, P. Ruiz-Flaño, and J. Castroviejo, 2015: Effects of farming
17 terraces on hydrological and geomorphological processes. A review. *CATENA*, **128**, 122–134,
18 doi:10.1016/J.CATENA.2015.01.021.
19 <https://www.sciencedirect.com/science/article/pii/S0341816215000351> (Accessed October 28,
20 2018).
- 21 Arnell, N. W., and S. N. Gosling, 2016: The impacts of climate change on river flood risk at the
22 global scale. *Clim. Change*, **134**, 387–401, doi:10.1007/s10584-014-1084-5.
23 <http://link.springer.com/10.1007/s10584-014-1084-5> (Accessed March 18, 2019).
- 24 Asmelash, F., T. Bekele, and E. Birhane, 2016: The Potential Role of Arbuscular Mycorrhizal Fungi
25 in the Restoration of Degraded Lands. *Front. Microbiol.*, **7**, 1095,
26 doi:10.3389/fmicb.2016.01095.
27 <http://journal.frontiersin.org/Article/10.3389/fmicb.2016.01095/abstract> (Accessed March 6,
28 2019).
- 29 Asner, G. P., A. J. Elmore, L. P. Olander, R. E. Martin, and A. T. Harris, 2004: Grazing Systems,
30 Ecosystem Responses, and Global Change. *Annu. Rev. Environ. Resour.*, **29**, 261–299,
31 doi:10.1146/annurev.energy.29.062403.102142.
32 <http://www.annualreviews.org/doi/10.1146/annurev.energy.29.062403.102142> (Accessed
33 December 28, 2017).
- 34 Astrup, R., P. Y. Bernier, H. Genet, D. A. Lutz, and R. M. Bright, 2018: A sensible climate solution
35 for the boreal forest. *Nat. Clim. Chang.*, doi:10.1038/s41558-017-0043-3.
- 36 Atmadja, S., and L. Verchot, 2012: A review of the state of research, policies and strategies in
37 addressing leakage from reducing emissions from deforestation and forest degradation
38 (REDD+). *Mitig. Adapt. Strateg. Glob. Chang.*, **17**, 311–336, doi:10.1007/s11027-011-9328-4.
39 <http://link.springer.com/10.1007/s11027-011-9328-4> (Accessed October 25, 2018).
- 40 Atwood, T. B., and Coauthors, 2017: Global patterns in mangrove soil carbon stocks and losses. *Nat.*
41 *Clim. Chang.*, **7**, 523–528, doi:10.1038/nclimate3326.
42 <http://www.nature.com/doi/10.1038/nclimate3326> (Accessed July 30, 2018).
- 43 Van Auken, O. W., 2009: Causes and consequences of woody plant encroachment into western North
44 American grasslands. *J. Environ. Manage.*, **90**, 2931–2942, doi:10.1016/j.jenvman.2009.04.023.
45 <https://www.sciencedirect.com/science/article/pii/S0301479709001522> (Accessed December 28,
46 2017).
- 47 Baccini, A., W. Walker, L. Carvalho, M. Farina, D. Sulla-Menashe, and R. A. Houghton, 2017:
48 Tropical forests are a net carbon source based on aboveground measurements of gain and loss.
49 *Science*, **358**, 230–234, doi:10.1126/science.aam5962.

- 1 <http://www.ncbi.nlm.nih.gov/pubmed/28971966> (Accessed May 26, 2018).
- 2 Bacmeister, J. T., K. A. Reed, C. Hannay, P. Lawrence, S. Bates, J. E. Truesdale, N. Rosenbloom, and
3 M. Levy, 2018: Projected changes in tropical cyclone activity under future warming scenarios
4 using a high-resolution climate model. *Clim. Change*, **146**, 547–560, doi:10.1007/s10584-016-
5 1750-x. <http://link.springer.com/10.1007/s10584-016-1750-x> (Accessed April 10, 2019).
- 6 Bae, J. S., R. W. Joo, Y.-S. Kim, and Y. S. Kim Yeon-Su., 2012: Forest transition in South Korea:
7 Reality, path and drivers. *Land use policy*, **29**, 198–207, doi:10.1016/j.landusepol.2011.06.007.
8 <https://www.sciencedirect.com/science/article/pii/S0264837711000615> (Accessed May 21,
9 2018).
- 10 Bahamondez, C., and I. D. Thompson, 2016: Determining forest degradation, ecosystem state and
11 resilience using a standard stand stocking measurement diagram: theory into practice. *Forestry*,
12 **89**, 290–300, doi:10.1093/forestry/cpv052. [https://academic.oup.com/forestry/article-](https://academic.oup.com/forestry/article-lookup/doi/10.1093/forestry/cpv052)
13 [lookup/doi/10.1093/forestry/cpv052](https://academic.oup.com/forestry/article-lookup/doi/10.1093/forestry/cpv052) (Accessed October 1, 2018).
- 14 Bai, Y., J. Wu, Q. Xing, Q. Pan, J. Huang, D. Yang, and X. Han, 2008a: PRIMARY PRODUCTION
15 AND RAIN USE EFFICIENCY ACROSS A PRECIPITATION GRADIENT ON THE
16 MONGOLIA PLATEAU. *Ecology*, **89**, 2140–2153, doi:10.1890/07-0992.1.
17 <http://doi.wiley.com/10.1890/07-0992.1> (Accessed January 3, 2018).
- 18 Bai, Z., D. Dent, L. Olsson, and M. E. Schaepman, 2008b: *Global Assessment of Land Degradation*
19 *and Improvment. 1. Identification by Remote Sensing*. Wageningen, The Netherlands, 78 pp.
- 20 Bai, Z. G., D. L. Dent, L. Olsson, and M. E. Schaepman, 2008c: Proxy global assessment of land
21 degradation. *Soil Use Manag.*, **24**, 223–234, doi:10.1111/j.1475-2743.2008.00169.x.
22 <http://doi.wiley.com/10.1111/j.1475-2743.2008.00169.x> (Accessed October 19, 2017).
- 23 Bai, Z. G., D. L. Dent, L. Olsson, A. Tengberg, C. J. Tucker, and G. Yengoh, 2015: A longer, closer,
24 look at land degradation. *Agric. Dev.*, **24**, 3–9. <http://library.wur.nl/WebQuery/wurpubs/490477>
25 (Accessed May 25, 2018).
- 26 Bailis, R., R. Drigo, A. Ghilardi, and O. Masera, 2015: The carbon footprint of traditional woodfuels.
27 *Nat. Clim. Chang.*, **5**, 266–272, doi:10.1038/nclimate2491.
28 <http://www.nature.com/articles/nclimate2491> (Accessed November 2, 2018).
- 29 Baka, J., 2013: The Political Construction of Wasteland: Governmentality, Land Acquisition and
30 Social Inequality in South India. *Dev. Change*, **44**, 409–428, doi:10.1111/dech.12018.
31 <http://doi.wiley.com/10.1111/dech.12018> (Accessed April 11, 2019).
- 32 —, 2014: What wastelands? A critique of biofuel policy discourse in South India. *Geoforum*, **54**,
33 315–323, doi:10.1016/J.GEOFORUM.2013.08.007.
34 <https://www.sciencedirect.com/science/article/pii/S0016718513001760> (Accessed April 11,
35 2019).
- 36 Baker, B., 2017: Can modern agriculture be sustainable? *Bioscience*, **67**, 325–331,
37 doi:10.1093/biosci/bix018.
- 38 Baker, J. M., T. E. Ochsner, R. T. Venterea, and T. J. Griffis, 2007: Tillage and soil carbon
39 sequestration—What do we really know? *Agric. Ecosyst. Environ.*, **118**, 1–5,
40 doi:10.1016/J.AGEE.2006.05.014.
41 <https://www.sciencedirect.com/science/article/pii/S0167880906001617> (Accessed March 31,
42 2019).
- 43 Bakun, A., 1990: Global climate change and intensification of coastal ocean upwelling. *Science*, **247**,
44 198–201, doi:10.1126/science.247.4939.198. <http://www.ncbi.nlm.nih.gov/pubmed/17813287>
45 (Accessed October 24, 2018).
- 46 Bakun, A., B. A. Black, S. J. Bograd, M. García-Reyes, A. J. Miller, R. R. Rykaczewski, and W. J.
47 Sydeman, 2015: Anticipated Effects of Climate Change on Coastal Upwelling Ecosystems.
48 *Curr. Clim. Chang. Reports*, **1**, 85–93, doi:10.1007/s40641-015-0008-4.

- 1 <http://link.springer.com/10.1007/s40641-015-0008-4> (Accessed March 8, 2019).
- 2 Bala, G., K. Caldeira, M. Wickett, T. J. Phillips, D. B. Lobell, C. Delire, and A. Mirin, 2007:
3 Combined climate and carbon-cycle effects of large-scale deforestation. *Proc. Natl. Acad. Sci.*,
4 **104**, 6550–6555, doi:10.1073/pnas.0608998104.
- 5 Balbo, A. L., 2017: Terrace landscapes. Editorial to the special issue. *J. Environ. Manage.*, **202**, 495–
6 499, doi:10.1016/J.JENVMAN.2017.02.001.
7 <https://www.sciencedirect.com/science/article/pii/S0301479717301056#bib1> (Accessed October
8 28, 2018).
- 9 Balooni, K., and J. F. Lund, 2014: Forest Rights: The Hard Currency of REDD+. *Conserv. Lett.*, **7**,
10 278–284, doi:10.1111/conl.12067. <http://doi.wiley.com/10.1111/conl.12067> (Accessed October
11 25, 2018).
- 12 Barbier, E. B., 2000: Valuing the environment as input: review of applications to mangrove-fishery
13 linkages. *Ecol. Econ.*, **35**, 47–61, doi:10.1016/S0921-8009(00)00167-1.
14 <https://www.sciencedirect.com/science/article/pii/S0921800900001671> (Accessed May 15,
15 2018).
- 16 Barbier, E. B., and J. P. Hochard, 2016: Does Land Degradation Increase Poverty in Developing
17 Countries? *PLoS One*, **11**, e0152973, doi:10.1371/journal.pone.0152973.
18 <https://dx.plos.org/10.1371/journal.pone.0152973> (Accessed February 13, 2019).
- 19 —, and —, 2018: Land degradation and poverty. *Nat. Sustain.*, **1**, 623–631, doi:10.1038/s41893-
20 018-0155-4. <http://www.nature.com/articles/s41893-018-0155-4> (Accessed March 5, 2019).
- 21 Bárcena, T. G., L. P. Kiaer, L. Vesterdal, H. M. Stefánsdóttir, P. Gundersen, and B. D. Sigurdsson,
22 2014: Soil carbon stock change following afforestation in Northern Europe: a meta-analysis.
23 *Glob. Chang. Biol.*, **20**, 2393–2405, doi:10.1111/gcb.12576.
24 <http://doi.wiley.com/10.1111/gcb.12576> (Accessed October 21, 2018).
- 25 Bardsley, D. K., and G. J. Hugo, 2010: Migration and climate change: examining thresholds of
26 change to guide effective adaptation decision-making. *Popul. Environ.*, **32**, 238–262,
27 doi:10.1007/s11111-010-0126-9. <http://link.springer.com/10.1007/s11111-010-0126-9>
28 (Accessed March 30, 2019).
- 29 Barlow, J., and Coauthors, 2007: Quantifying the biodiversity value of tropical primary, secondary,
30 and plantation forests. *Proc. Natl. Acad. Sci. U. S. A.*, **104**, 18555–18560,
31 doi:10.1073/pnas.0703333104. <http://www.ncbi.nlm.nih.gov/pubmed/18003934> (Accessed May
32 26, 2018).
- 33 Barnett, J., and S. O’Neill, 2010: Maladaptation. *Glob. Environ. Chang.*, **20**, 211–213,
34 doi:10.1016/j.gloenvcha.2009.11.004.
- 35 —, L. S. Evans, C. Gross, A. S. Kiem, R. T. Kingsford, J. P. Palutikof, C. M. Pickering, and S. G.
36 Smithers, 2015: From barriers to limits to climate change adaptation: path dependency and the
37 speed of change. *Ecol. Soc.*, **20**, art5, doi:10.5751/ES-07698-200305.
38 <http://www.ecologyandsociety.org/vol20/iss3/art5/> (Accessed March 29, 2019).
- 39 Barnhart, T. B., N. P. Molotch, B. Livneh, A. A. Harpold, J. F. Knowles, and D. Schneider, 2016:
40 Snowmelt rate dictates streamflow. *Geophys. Res. Lett.*, **43**, 8006–8016,
41 doi:10.1002/2016GL069690. <http://doi.wiley.com/10.1002/2016GL069690> (Accessed March 6,
42 2019).
- 43 Barrera-Bassols, N., and J. A. Zinck, 2003: Ethnopedology: a worldwide view on the soil knowledge
44 of local people. *Geoderma*, **111**, 171–195, doi:10.1016/S0016-7061(02)00263-X.
45 <https://www.sciencedirect.com/science/article/pii/S001670610200263X> (Accessed October 28,
46 2018).
- 47 Barring, L., P. Jönsson, J. O. Mattsson, and R. Åhman, 2003: Wind erosion on arable land in Scania,
48 Sweden and the relation to the wind climate—a review. *CATENA*, **52**, 173–190,

- 1 doi:10.1016/S0341-8162(03)00013-4.
2 <https://www.sciencedirect.com/science/article/pii/S0341816203000134> (Accessed February 14,
3 2018).
- 4 Basche, A., and M. DeLonge, 2017: The Impact of Continuous Living Cover on Soil Hydrologic
5 Properties: A Meta-Analysis. *Soil Sci. Soc. Am. J.*, **81**, 1179, doi:10.2136/sssaj2017.03.0077.
6 <https://dl.sciencesocieties.org/publications/sssaj/abstracts/81/5/1179> (Accessed October 18,
7 2018).
- 8 Batello, C., L. Wade, S. Cox, N. Pogna, A. Bozzini, and J. Choptiany, 2014: *Perennial crops for food*
9 *security*. <http://agris.fao.org/agris-search/search.do?recordID=XF2017002349> (Accessed
10 October 31, 2017).
- 11 Batir, J. F., M. J. Hornbach, and D. D. Blackwell, 2017: Ten years of measurements and modeling of
12 soil temperature changes and their effects on permafrost in Northwestern Alaska. *Glob. Planet.*
13 *Change*, **148**, 55–71, doi:10.1016/J.GLOPLACHA.2016.11.009.
14 <https://www.sciencedirect.com/science/article/pii/S0921818116303034> (Accessed March 16,
15 2019).
- 16 Bauer, N., and Coauthors, 2018: Global energy sector emission reductions and bioenergy use:
17 overview of the bioenergy demand phase of the EMF-33 model comparison. *Clim. Change*, 1–
18 16, doi:10.1007/s10584-018-2226-y. <http://link.springer.com/10.1007/s10584-018-2226-y>
19 (Accessed April 11, 2019).
- 20 Baveye, P. C., 2017: Quantification of ecosystem services: Beyond all the “guesstimates”, how do we
21 get real data? *Ecosyst. Serv.*, **24**, 47–49, doi:10.1016/J.ECOSER.2017.02.006.
22 <https://www.sciencedirect.com/science/article/pii/S2212041616302856?via%3Dihub> (Accessed
23 February 20, 2019).
- 24 ——, D. Rangel, A. R. Jacobson, M. Laba, C. Darnault, W. Otten, R. Radulovich, and F. A. O.
25 Camargo, 2011: From Dust Bowl to Dust Bowl: Soils are Still Very Much a Frontier of Science.
26 *Soil Sci. Soc. Am. J.*, **75**, 2037, doi:10.2136/sssaj2011.0145.
27 <https://www.soils.org/publications/sssaj/abstracts/75/6/2037> (Accessed October 30, 2018).
- 28 Beck, P. S. A., and Coauthors, 2011: Changes in forest productivity across Alaska consistent with
29 biome shift. *Ecol. Lett.*, **14**, 373–379, doi:10.1111/j.1461-0248.2011.01598.x.
30 <http://doi.wiley.com/10.1111/j.1461-0248.2011.01598.x> (Accessed October 27, 2018).
- 31 Beerling, D. J., 2017: Enhanced rock weathering: biological climate change mitigation with co-
32 benefits for food security? *Biol. Lett.*, **13**, 20170149, doi:10.1098/rsbl.2017.0149.
33 <http://rsbl.royalsocietypublishing.org/lookup/doi/10.1098/rsbl.2017.0149> (Accessed April 12,
34 2019).
- 35 Beets, W. C., 1990: *Raising and sustaining productivity of smallholder farming systems in the tropics:*
36 *a handbook of sustainable agricultural development*. Agbe Publishing, Alkmaar, The
37 Netherlands, 754 pp. <https://www.cabdirect.org/cabdirect/abstract/19916780980> (Accessed
38 October 30, 2018).
- 39 Behnke, R., 1994: Natural Resource Management in Pastoral Africa. *Dev. Policy Rev.*, **12**, 5–28,
40 doi:10.1111/j.1467-7679.1994.tb00053.x.
- 41 ——, and M. Mortimore, 2016: Introduction: The End of Desertification? Springer, Berlin,
42 Heidelberg, 1–34 http://link.springer.com/10.1007/978-3-642-16014-1_1 (Accessed May 12,
43 2018).
- 44 Belair, E. P., and M. J. Ducey, 2018: Patterns in Forest Harvesting in New England and New York:
45 Using FIA Data to Evaluate Silvicultural Outcomes. *J. For.*, **116**, 273–282,
46 doi:10.1093/jofore/fvx019. <https://academic.oup.com/jof/article/116/3/273/4993098> (Accessed
47 May 31, 2018).
- 48 Bellamy, P. H., P. J. Loveland, R. I. Bradley, R. M. Lark, and G. J. D. Kirk, 2005: Carbon losses from
49 all soils across England and Wales 1978–2003. *Nature*, **437**, 245–248, doi:10.1038/nature04038.

- 1 <http://www.nature.com/articles/nature04038> (Accessed November 2, 2018).
- 2 Belnap, J., B. J. Walker, S. M. Munson, and R. A. Gill, 2014: Controls on sediment production in two
3 U.S. deserts. *Aeolian Res.*, **14**, 15–24, doi:10.1016/J.AEOLIA.2014.03.007.
4 <https://www.sciencedirect.com/science/article/pii/S1875963714000263> (Accessed October 2,
5 2018).
- 6 Benavidez, R., B. Jackson, D. Maxwell, and K. Norton, 2018: A review of the (Revised) Universal
7 Soil Loss Equation (R/USLE): with a view to increasing its global applicability and improving
8 soil loss estimates. *Hydrol. Earth Syst. Sci. Discuss.*, 1–34, doi:10.5194/hess-2018-68.
9 <https://doi.org/10.5194/hess-2018-68> (Accessed October 1, 2018).
- 10 Bender, M. A., T. R. Knutson, R. E. Tuleya, J. J. Sirutis, G. A. Vecchi, S. T. Garner, and I. M. Held,
11 2010: Modeled Impact of Anthropogenic Warming on the Frequency of Intense Atlantic
12 Hurricanes. *Science* (80-.), **327**, 454 LP-458.
13 <http://science.sciencemag.org/content/327/5964/454.abstract>.
- 14 Béné, C., R. G. Wood, A. Newsham, and M. Davies, 2012: Resilience: New Utopia or New Tyranny?
15 Reflection about the Potentials and Limits of the Concept of Resilience in Relation to
16 Vulnerability Reduction Programmes. *IDS Work. Pap.*, **2012**, 1–61, doi:10.1111/j.2040-
17 0209.2012.00405.x. <http://doi.wiley.com/10.1111/j.2040-0209.2012.00405.x> (Accessed May 31,
18 2018).
- 19 Benini, L., M. Antonellini, M. Laghi, and P. N. Mollema, 2016: Assessment of Water Resources
20 Availability and Groundwater Salinization in Future Climate and Land use Change Scenarios: A
21 Case Study from a Coastal Drainage Basin in Italy. *Water Resour. Manag.*, **30**, 731–745,
22 doi:10.1007/s11269-015-1187-4. <http://link.springer.com/10.1007/s11269-015-1187-4>
23 (Accessed March 18, 2019).
- 24 Benjaminsen, T. A., and C. Lund, 2003: *Securing land rights in Africa*. Frank Cass Publishers,
25 London, UK, 175 pp. <https://www.cabdirect.org/cabdirect/abstract/20053122959> (Accessed May
26 16, 2018).
- 27 Benjaminsen, T. A., K. Alinon, H. Buhaug, and J. T. Buseeth, 2012: Does climate change drive land-
28 use conflicts in the Sahel? *J. Peace Res.*, **49**, 97–111, doi:10.1177/0022343311427343.
29 <http://journals.sagepub.com/doi/10.1177/0022343311427343> (Accessed May 15, 2018).
- 30 Bennett, M. T., 2008: China’s sloping land conversion program: Institutional innovation or business
31 as usual? *Ecol. Econ.*, doi:10.1016/j.ecolecon.2007.09.017.
- 32 Bentz, B. J., and Coauthors, 2010: Climate Change and Bark Beetles of the Western United States and
33 Canada: Direct and Indirect Effects. *Bioscience*, **60**, 602–613, doi:10.1525/bio.2010.60.8.6.
34 <https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2010.60.8.6> (Accessed
35 March 5, 2018).
- 36 Bernal, B., L. T. Murray, and T. R. H. Pearson, 2018: Global carbon dioxide removal rates from forest
37 landscape restoration activities. *Carbon Balance Manag.*, **13**, doi:10.1186/s13021-018-0110-8.
38 <https://doi.org/10.1186/s13021-018-0110-8>.
- 39 Bernier, P. Y., and Coauthors, 2017: Moving beyond the concept of “primary forest” as a metric of
40 forest environment quality. *Ecol. Appl.*, **27**, 349–354, doi:10.1002/eap.1477.
41 <http://doi.wiley.com/10.1002/eap.1477> (Accessed May 26, 2018).
- 42 Bernoux, M., B. Volkoff, M. da C. S. Carvalho, and C. C. Cerri, 2003: CO₂ emissions from liming of
43 agricultural soils in Brazil. *Global Biogeochem. Cycles*, **17**, n/a–n/a,
44 doi:10.1029/2001GB001848. <http://doi.wiley.com/10.1029/2001GB001848> (Accessed May 8,
45 2018).
- 46 Betts, R. A., 2000: Offset of the potential carbon sink from boreal forestation by decreases in surface
47 albedo. *Nature*, **408**, 187–190, doi:10.1038/35041545. <http://www.nature.com/articles/35041545>
48 (Accessed December 26, 2017).

- 1 —, and Coauthors, 2007: Projected increase in continental runoff due to plant responses to
2 increasing carbon dioxide. *Nature*, **448**, 1037–1041, doi:10.1038/nature06045.
- 3 Betzold, C., and I. Mohamed, 2017: Seawalls as a response to coastal erosion and flooding: a case
4 study from Grande Comore, Comoros (West Indian Ocean). *Reg. Environ. Chang.*, **17**, 1077–
5 1087, doi:10.1007/s10113-016-1044-x.
- 6 Bharucha, Z., and J. Pretty, 2010: The roles and values of wild foods in agricultural systems. *Philos.*
7 *Trans. R. Soc. B Biol. Sci.*, **365**, 2913–2926, doi:10.1098/rstb.2010.0123.
8 <http://www.royalsocietypublishing.org/doi/10.1098/rstb.2010.0123> (Accessed March 8, 2019).
- 9 Bhatia, K., and Coauthors, 2018: Projected Response of Tropical Cyclone Intensity and
10 Intensification in a Global Climate Model. *J. Clim.*, **31**, 8281–8303, doi:10.1175/JCLI-D-17-
11 0898.1. <http://journals.ametsoc.org/doi/10.1175/JCLI-D-17-0898.1> (Accessed October 14,
12 2018).
- 13 Biederman, L. A., and W. Stanley Harpole, 2013: Biochar and its effects on plant productivity and
14 nutrient cycling: A meta-analysis. *GCB Bioenergy*, doi:10.1111/gcbb.12037.
- 15 Bindoff, N. L., and Coauthors, 2013: Detection and Attribution of Climate Change: from Global to
16 Regional. *Climate Change 2013: The Physical Science Basis. Contribution of Working Group I*
17 *to the Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, T.F. Stocker
18 et al., Eds., Cambridge University Press, Cambridge, UK and New York, USA, 867–940.
- 19 Birdsey, R., K. Pregitzer, and A. Lucier, 2006: Forest Carbon Management in the United States. *J.*
20 *Environ. Qual.*, **35**, 1461, doi:10.2134/jeq2005.0162.
21 <https://www.agronomy.org/publications/jeq/abstracts/35/4/1461> (Accessed May 26, 2018).
- 22 Bisaro, A., M. Kirk, P. Zdruli, and W. Zimmermann, 2014: GLOBAL DRIVERS SETTING
23 DESERTIFICATION RESEARCH PRIORITIES: INSIGHTS FROM A STAKEHOLDER
24 CONSULTATION FORUM. *L. Degrad. Dev.*, **25**, 5–16, doi:10.1002/ldr.2220.
25 <http://doi.wiley.com/10.1002/ldr.2220> (Accessed March 1, 2019).
- 26 Bitzer, V., and J. Bijman, 2015: From innovation to co-innovation? An exploration of African
27 agrifood chains. *Br. Food J.*, **117**, 2182–2199, doi:10.1108/BFJ-12-2014-0403.
28 <http://www.emeraldinsight.com/doi/10.1108/BFJ-12-2014-0403> (Accessed March 25, 2019).
- 29 Black, R., S. R. G. Bennett, S. M. Thomas, and J. R. Beddington, 2011: Climate change: Migration as
30 adaptation. *Nat. 2011 4787370*.
- 31 Blaikie, P. M., and H. C. Brookfield, 1987: *Land degradation and society*. P.M. Blaikie and H.C.
32 Brookfield, Eds. Methuen, Milton Park, Abingdon, UK.
- 33 Blanco-Canqui, H., and R. Lal, 2009: Crop Residue Removal Impacts on Soil Productivity and
34 Environmental Quality. *CRC. Crit. Rev. Plant Sci.*, **28**, 139–163,
35 doi:10.1080/07352680902776507. <https://doi.org/10.1080/07352680902776507>.
- 36 Blaser, W. J., J. Oppong, E. Yeboah, and J. Six, 2017: Shade trees have limited benefits for soil
37 fertility in cocoa agroforests. *Agric. Ecosyst. Environ.*, **243**, 83–91,
38 doi:10.1016/J.AGEE.2017.04.007.
39 <https://www.sciencedirect.com/science/article/pii/S0167880917301615> (Accessed November 1,
40 2018).
- 41 Blenkinsop, S., and Coauthors, 2018: The INTENSE project: using observations and models to
42 understand the past, present and future of sub-daily rainfall extremes. *Adv. Sci. Res.*, **15**, 117–
43 126, doi:10.5194/asr-15-117-2018. <https://www.adv-sci-res.net/15/117/2018/> (Accessed October
44 22, 2018).
- 45 Boardman, J., 2010: A short history of muddy floods. *L. Degrad. Dev.*, **21**, 303–309,
46 doi:10.1002/ldr.1007. <http://doi.wiley.com/10.1002/ldr.1007> (Accessed March 15, 2018).
- 47 Bocquet-Appel, J.-P., 2011: When the world’s population took off: the springboard of the Neolithic
48 Demographic Transition. *Science*, **333**, 560–561, doi:10.1126/science.1208880.

- 1 <http://www.ncbi.nlm.nih.gov/pubmed/21798934> (Accessed February 27, 2019).
- 2 Boit, A., and Coauthors, 2016: Large-scale impact of climate change vs. land-use change on future
3 biome shifts in Latin America. *Glob. Chang. Biol.*, **22**, 3689–3701, doi:10.1111/gcb.13355.
4 <http://doi.wiley.com/10.1111/gcb.13355> (Accessed March 14, 2019).
- 5 Bonan, G. B., and Coauthors, 2008: Forests and climate change: forcings, feedbacks, and the climate
6 benefits of forests. *Science*, **320**, 1444–1449, doi:10.1126/science.1155121.
7 <http://www.ncbi.nlm.nih.gov/pubmed/18556546> (Accessed March 5, 2018).
- 8 Bonanomi, G., F. Ippolito, G. Cesarano, F. Vinale, N. Lombardi, A. Crasto, S. L. Woo, and F. Scala,
9 2018: Biochar chemistry defined by ¹³C-CPMAS NMR explains opposite effects on soilborne
10 microbes and crop plants. *Appl. Soil Ecol.*, doi:10.1016/j.apsoil.2017.11.027.
- 11 Bond-Lamberty, B., V. L. Bailey, M. Chen, C. M. Gough, and R. Vargas, 2018: Globally rising soil
12 heterotrophic respiration over recent decades. *Nature*, **560**, 80–83, doi:10.1038/s41586-018-
13 0358-x. <http://www.nature.com/articles/s41586-018-0358-x> (Accessed October 1, 2018).
- 14 Bond, T. C., and Coauthors, 2013: Bounding the role of black carbon in the climate system: A
15 scientific assessment. *J. Geophys. Res. Atmos.*, **118**, 5380–5552, doi:10.1002/jgrd.50171.
16 <https://doi.org/10.1002/jgrd.50171>.
- 17 Bonn, A., Allott, T., Evans, M., Joosten, H., Stoneman, R., 2016: Peatland restoration and ecosystem
18 services: nature-based solutions for societal goals. In: Bonn A., Allott T., Evans M., Joosten H.,
19 Stoneman R., (Eds.), Peatland restoration and ecosystem services: science, policy and practice.
20 Cambridge University Pres. 402–417.
- 21 Borchard, N., and Coauthors, 2019: Biochar, soil and land-use interactions that reduce nitrate leaching
22 and N₂O emissions: A meta-analysis. *Sci. Total Environ.*, **651**, 2354–2364,
23 doi:10.1016/J.SCITOTENV.2018.10.060.
24 <https://www.sciencedirect.com/science/article/pii/S0048969718339330> (Accessed April 2,
25 2019).
- 26 Borrás, S. M., and J. C. Franco, 2018: The challenge of locating land-based climate change mitigation
27 and adaptation politics within a social justice perspective: towards an idea of agrarian climate
28 justice. *Third World Q.*, **39**, 1308–1325, doi:10.1080/01436597.2018.1460592.
29 <https://www.tandfonline.com/doi/full/10.1080/01436597.2018.1460592> (Accessed April 12,
30 2019).
- 31 Borrelli, P., K. Paustian, P. Panagos, A. Jones, B. Schütt, and E. Lugato, 2016: Effect of Good
32 Agricultural and Environmental Conditions on erosion and soil organic carbon balance: A
33 national case study. *Land use policy*, **50**, 408–421, doi:10.1016/J.LANDUSEPOL.2015.09.033.
34 <https://www.sciencedirect.com/science/article/pii/S0264837715003257> (Accessed April 3,
35 2019).
- 36 ———, and Coauthors, 2017: An assessment of the global impact of 21st century land use change on
37 soil erosion. *Nat. Commun.*, **8**, 2013, doi:10.1038/s41467-017-02142-7.
38 <http://www.nature.com/articles/s41467-017-02142-7> (Accessed October 25, 2018).
- 39 Borrelli, P., K. Van Oost, K. Meusburger, C. Alewell, E. Lugato, and P. Panagos, 2018: A step
40 towards a holistic assessment of soil degradation in Europe: Coupling on-site erosion with
41 sediment transfer and carbon fluxes. *Environ. Res.*, **161**, 291–298,
42 doi:10.1016/J.ENVRES.2017.11.009.
43 <https://www.sciencedirect.com/science/article/pii/S0013935117308137> (Accessed March 18,
44 2019).
- 45 Bosire, J. O., F. Dahdouh-Guebas, M. Walton, B. I. Crona, R. R. Lewis, C. Field, J. G. Kairo, and N.
46 Koedam, 2008: Functionality of restored mangroves: A review. *Aquat. Bot.*, **89**, 251–259,
47 doi:10.1016/J.AQUABOT.2008.03.010.
48 <https://www.sciencedirect.com/science/article/pii/S0304377008000521> (Accessed July 30,
49 2018).

- 1 Bot, A., F. Nachtergaele, and A. Young, 2000: *Land resource potential and constraints at regional*
2 *and country levels*. Land and Water Development Division, Food and Agriculture Organization
3 of the United Nations, Rome, Italy, 114 pp.
- 4 Bousbih, S., and Coauthors, 2017: Potential of Sentinel-1 Radar Data for the Assessment of Soil and
5 Cereal Cover Parameters. *Sensors*, **17**, 2617, doi:10.3390/s17112617.
6 <http://www.mdpi.com/1424-8220/17/11/2617> (Accessed October 1, 2018).
- 7 Bowles, T. M., S. S. Atallah, E. E. Campbell, A. C. M. Gaudin, W. R. Wieder, and A. S. Grandy,
8 2018: Addressing agricultural nitrogen losses in a changing climate. *Nat. Sustain.*, **1**, 399–408,
9 doi:10.1038/s41893-018-0106-0. <http://www.nature.com/articles/s41893-018-0106-0> (Accessed
10 October 18, 2018).
- 11 Bozzi, E., L. Genesio, P. Toscano, M. Pieri, and F. Miglietta, 2015: Mimicking biochar-albedo
12 feedback in complex Mediterranean agricultural landscapes. *Environ. Res. Lett.*,
13 doi:10.1088/1748-9326/10/8/084014.
- 14 Bradshaw, C. J. A., N. S. Sodhi, K. S.-H. Peh, and B. W. Brook, 2007: Global evidence that
15 deforestation amplifies flood risk and severity in the developing world. *Glob. Chang. Biol.*, **13**,
16 2379–2395, doi:10.1111/j.1365-2486.2007.01446.x. <http://doi.wiley.com/10.1111/j.1365->
17 [2486.2007.01446.x](http://doi.wiley.com/10.1111/j.1365-2486.2007.01446.x) (Accessed December 27, 2017).
- 18 Brahney, J., F. Weber, V. Foord, J. Janmaat, and P. J. Curtis, 2017: Evidence for a climate-driven
19 hydrologic regime shift in the Canadian Columbia Basin. *Can. Water Resour. J. / Rev. Can. des*
20 *ressources hydriques*, **42**, 179–192, doi:10.1080/07011784.2016.1268933.
21 <https://www.tandfonline.com/doi/full/10.1080/07011784.2016.1268933> (Accessed March 6,
22 2019).
- 23 Branca, G., L. Lipper, N. McCarthy, and M. C. Jolejole, 2013: Food security, climate change, and
24 sustainable land management. A review. *Agron. Sustain. Dev.*, **33**, 635–650,
25 doi:10.1007/s13593-013-0133-1.
- 26 Brandt, M., and Coauthors, 2017: Human population growth offsets climate-driven increase in woody
27 vegetation in sub-Saharan Africa. *Nat. Ecol. Evol.*, **1**, 0081, doi:10.1038/s41559-017-0081.
28 <http://www.nature.com/articles/s41559-017-0081> (Accessed May 23, 2018).
- 29 —, and Coauthors, 2018a: Reduction of tree cover in West African woodlands and promotion in
30 semi-arid farmlands. *Nat. Geosci.*, **11**, 328–333, doi:10.1038/s41561-018-0092-x.
31 <http://www.nature.com/articles/s41561-018-0092-x> (Accessed May 23, 2018).
- 32 —, and Coauthors, 2018b: Satellite passive microwaves reveal recent climate-induced carbon
33 losses in African drylands. *Nat. Ecol. Evol.*, **2**, 827–835, doi:10.1038/s41559-018-0530-6.
34 <http://www.nature.com/articles/s41559-018-0530-6> (Accessed May 23, 2018).
- 35 Bright, R. M., K. Zhao, R. B. Jackson, and F. Cherubini, 2015: Quantifying surface albedo and other
36 direct biogeophysical climate forcings of forestry activities. *Glob. Chang. Biol.*, **21**, 3246–3266,
37 doi:10.1111/gcb.12951. <http://doi.wiley.com/10.1111/gcb.12951> (Accessed May 12, 2018).
- 38 Brinkman, E. P., R. Postma, W. H. van der Putten, and A. J. Termorshuizen, 2017: Influence of
39 growing Eucalyptus trees for biomass on soil quality. 1–39.
40 [https://www.narcis.nl/publication/RecordID/oai:pure.knaw.nl:publications%2F15ebd62c-9252-](https://www.narcis.nl/publication/RecordID/oai:pure.knaw.nl:publications%2F15ebd62c-9252-40d6-b54e-12f1d592ecbc)
41 [40d6-b54e-12f1d592ecbc](https://www.narcis.nl/publication/RecordID/oai:pure.knaw.nl:publications%2F15ebd62c-9252-40d6-b54e-12f1d592ecbc) (Accessed April 11, 2019).
- 42 Brisson, N., P. Gate, D. Gouache, G. Charmet, F.-X. Oury, and F. Huard, 2010: Why are wheat yields
43 stagnating in Europe? A comprehensive data analysis for France. *F. Crop. Res.*, **119**, 201–212,
44 doi:10.1016/J.FCR.2010.07.012.
45 <https://www.sciencedirect.com/science/article/pii/S0378429010001929> (Accessed May 23,
46 2018).
- 47 Brockington, J. D., I. M. Harris, and R. M. Brook, 2016: Beyond the project cycle: a medium-term
48 evaluation of agroforestry adoption and diffusion in a south Indian village. *Agrofor. Syst.*, **90**,
49 489–508, doi:10.1007/s10457-015-9872-0. <http://link.springer.com/10.1007/s10457-015-9872-0>

- 1 (Accessed October 22, 2018).
- 2 Brooks, M. L., and Coauthors, 2004: Effects of Invasive Alien Plants on Fire Regimes. *Bioscience*,
3 **54**, 677–688, doi:10.1641/0006-3568(2004)054[0677:eoiapo]2.0.co;2.
4 <https://academic.oup.com/bioscience/article/54/7/677/223532> (Accessed October 1, 2018).
- 5 Brooks, M. L., and Coauthors, 2009: Effects of Invasive Alien Plants on Fire Regimes.
6 [http://dx.doi.org/10.1641/0006-3568\(2004\)054\[0677:EOIAPO\]2.0.CO;2](http://dx.doi.org/10.1641/0006-3568(2004)054[0677:EOIAPO]2.0.CO;2), doi:10.1641/0006-
7 3568(2004)054[0677:EOIAPO]2.0.CO;2. [http://www.bioone.org/doi/abs/10.1641/0006-
8 3568\(2004\)054\[0677:EOIAPO\]2.0.CO;2](http://www.bioone.org/doi/abs/10.1641/0006-3568(2004)054[0677:EOIAPO]2.0.CO;2) (Accessed May 8, 2018).
- 9 Brown, K., 2014: Global environmental change I: A social turn for resilience? *Prog. Hum. Geogr.*, **38**,
10 107–117, doi:10.1177/0309132513498837.
11 <http://journals.sagepub.com/doi/10.1177/0309132513498837> (Accessed March 30, 2019).
- 12 Brown, S., and R. J. Nicholls, 2015: Subsidence and human influences in mega deltas: The case of the
13 Ganges–Brahmaputra–Meghna. *Sci. Total Environ.*, **527–528**, 362–374,
14 doi:10.1016/J.SCITOTENV.2015.04.124.
15 <https://www.sciencedirect.com/science/article/pii/S0048969715300589> (Accessed October 27,
16 2018).
- 17 Brunner, K.-M., S. Mandl, and H. Thomson, 2018: Energy Poverty: Energy equity in a world of high
18 demand and low supply. *The Oxford Handbook of Energy and Society*, D.J. Davidson and M.
19 Gross, Eds., Oxford University Press, New York, United States, 297–316
20 [https://books.google.se/books?hl=sv&lr=&id=eiRjDwAAQBAJ&oi=fnd&pg=PA297&dq=biom
21 ass+cooking+time+poverty&ots=m5XhMqzIXq&sig=pcuysjlq-
22 oUMYn5tH2PEAMAOn58&redir_esc=y#v=onepage&q=biomass_cooking_time
23 poverty&f=false](https://books.google.se/books?hl=sv&lr=&id=eiRjDwAAQBAJ&oi=fnd&pg=PA297&dq=biomass+cooking+time+poverty&ots=m5XhMqzIXq&sig=pcuysjlq-oUMYn5tH2PEAMAOn58&redir_esc=y#v=onepage&q=biomass_cooking_time_poverty&f=false) (Accessed April 10, 2019).
- 24 Buendia, C., R. J. Batalla, S. Sabater, A. Palau, and R. Marcé, 2016: Runoff Trends Driven by
25 Climate and Afforestation in a Pyrenean Basin. *L. Degrad. Dev.*, **27**, 823–838,
26 doi:10.1002/ldr.2384. <http://doi.wiley.com/10.1002/ldr.2384> (Accessed October 21, 2018).
- 27 Buggy, L., and K. E. McNamara, 2016: The need to reinterpret “community” for climate change
28 adaptation: a case study of Pele Island, Vanuatu. *Clim. Dev.*, **8**, 270–280,
29 doi:10.1080/17565529.2015.1041445.
- 30 Bunn, C., P. Läderach, O. Ovalle Rivera, and D. Kirschke, 2015: A bitter cup: climate change profile
31 of global production of Arabica and Robusta coffee. *Clim. Change*, **129**, 89–101,
32 doi:10.1007/s10584-014-1306-x. <http://link.springer.com/10.1007/s10584-014-1306-x>
33 (Accessed May 23, 2018).
- 34 Bürgi, M., and Coauthors, 2017: Integrated Landscape Approach: Closing the Gap between Theory
35 and Application. *Sustainability*, **9**, 1371, doi:10.3390/su9081371. [http://www.mdpi.com/2071-
36 1050/9/8/1371](http://www.mdpi.com/2071-1050/9/8/1371) (Accessed March 15, 2019).
- 37 Burkett, V., and J. Kusler, 2000: Climate change: potential impacts and interactions in wetlands in the
38 United States. *JAWRA J. Am. Water Resour. Assoc.*, **36**, 313–320, doi:10.1111/j.1752-
39 1688.2000.tb04270.x. <http://doi.wiley.com/10.1111/j.1752-1688.2000.tb04270.x> (Accessed
40 March 18, 2019).
- 41 Burlakova, L. E., A. Y. Karatayev, D. K. Padilla, L. D. Cartwright, and D. N. Hollas, 2009: Wetland
42 restoration and invasive species: Apple snail (*Pomacea insularum*) feeding on native and
43 invasive aquatic plants. *Restor. Ecol.*, **17**, 433–440, doi:10.1111/j.1526-100X.2008.00429.x.
44 <http://doi.wiley.com/10.1111/j.1526-100X.2008.00429.x> (Accessed July 18, 2018).
- 45 Burt, T., J. Boardman, I. Foster, and N. Howden, 2016a: More rain, less soil: long-term changes in
46 rainfall intensity with climate change. *Earth Surf. Process. Landforms*, **41**, 563–566,
47 doi:10.1002/esp.3868. <http://doi.wiley.com/10.1002/esp.3868> (Accessed December 5, 2017).
- 48 —, —, —, and —, 2016b: More rain, less soil: long-term changes in rainfall intensity with
49 climate change. *Earth Surf. Process. Landforms*, **41**, 563–566, doi:10.1002/esp.3868.

- 1 Busch, J., and Coauthors, 2015: Reductions in emissions from deforestation from Indonesia's
2 moratorium on new oil palm, timber, and logging concessions. *Proc. Natl. Acad. Sci. USA*, **112**,
3 1328–1333, doi:10.7910/DVN/28615. <http://www.pnas.org/content/112/5/1328.short>.
- 4 Bush, D. M., 2004: *Living with Florida's Atlantic beaches : coastal hazards from Amelia Island to*
5 *Key West*. Duke University Press, 338 pp.
- 6 Buss, W., O. Mašek, M. Graham, and D. Wüst, 2015: Inherent organic compounds in biochar-Their
7 content, composition and potential toxic effects. *J. Environ. Manage.*,
8 doi:10.1016/j.jenvman.2015.03.035.
- 9 Butler, R. A., L. P. Koh, and J. Ghazoul, 2009: REDD in the red: palm oil could undermine carbon
10 payment schemes. *Conserv. Lett.*, **2**, 67–73, doi:10.1111/j.1755-263X.2009.00047.x.
- 11 Butzer, K. W., 2005: Environmental history in the Mediterranean world: cross-disciplinary
12 investigation of cause-and-effect for degradation and soil erosion. *J. Archaeol. Sci.*, **32**, 1773–
13 1800, doi:10.1016/J.JAS.2005.06.001.
14 <https://www.sciencedirect.com/science/article/pii/S0305440305001275> (Accessed February 27,
15 2019).
- 16 Byers, M., and N. Dragojlovic, 2004: Darfur: A climate change-induced humanitarian crisis? *Hum.*
17 *Secur. Bull.*, 16–18.
- 18 Cacho, J. F., M. C. Negri, C. R. Zumpf, and P. Campbell, 2018: Introducing perennial biomass crops
19 into agricultural landscapes to address water quality challenges and provide other environmental
20 services. *Wiley Interdiscip. Rev. Energy Environ.*, **7**, e275, doi:10.1002/wene.275.
21 <http://doi.wiley.com/10.1002/wene.275> (Accessed April 11, 2019).
- 22 Cai, T., D. T. Price, A. L. Orchansky, and B. R. Thomas, 2011a: Carbon, Water, and Energy
23 Exchanges of a Hybrid Poplar Plantation During the First Five Years Following Planting.
24 *Ecosystems*, **14**, 658–671, doi:10.1007/s10021-011-9436-8.
- 25 Cai, X., X. Zhang, and D. Wang, 2011b: Land Availability for Biofuel Production. *Environ. Sci.*
26 *Technol.*, **45**, 334–339, doi:10.1021/es103338e. <http://pubs.acs.org/doi/abs/10.1021/es103338e>
27 (Accessed May 25, 2018).
- 28 Caldwell, P. V., C. R. Jackson, C. F. Miniati, S. E. Younger, J. A. Vining, J. J. McDonnell, and D. P.
29 Aubrey, 2018: Woody bioenergy crop selection can have large effects on water yield: A
30 southeastern United States case study. *Biomass and Bioenergy*, **117**, 180–189,
31 doi:10.1016/J.BIOMBIOE.2018.07.021.
32 <https://www.sciencedirect.com/science/article/pii/S0961953418301910> (Accessed April 11,
33 2019).
- 34 Camacho, L. D., D. T. Gevaña, †Antonio P. Carandang, and S. C. Camacho, 2016: Indigenous
35 knowledge and practices for the sustainable management of Ifugao forests in Cordillera,
36 Philippines. *Int. J. Biodivers. Sci. Ecosyst. Serv. Manag.*, **12**, 5–13,
37 doi:10.1080/21513732.2015.1124453.
38 <http://www.tandfonline.com/doi/full/10.1080/21513732.2015.1124453> (Accessed October 28,
39 2018).
- 40 Campioli, M., and Coauthors, 2015: Biomass production efficiency controlled by management in
41 temperate and boreal ecosystems. *Nat. Geosci.*, **8**, 843–846, doi:10.1038/ngeo2553.
- 42 Capolongo, D., N. Diodato, C. M. Mannaerts, M. Piccarreta, and R. O. Strobl, 2008: Analyzing
43 temporal changes in climate erosivity using a simplified rainfall erosivity model in Basilicata
44 (southern Italy). *J. Hydrol.*, **356**, 119–130, doi:10.1016/J.JHYDROL.2008.04.002.
45 <http://www.sciencedirect.com/science/article/pii/S0022169408001753> (Accessed December 5,
46 2017).
- 47 Carlson, K. M., and Coauthors, 2017: Effect of oil palm sustainability certification on deforestation
48 and fire in Indonesia. *Proc. Natl. Acad. Sci.*, **115**, 201704728, doi:10.1073/pnas.1704728114.

- 1 Carton, W., and E. Andersson, 2017: Where Forest Carbon Meets Its Maker: Forestry-Based
2 Offsetting as the Subsumption of Nature. *Soc. Nat. Resour.*, **30**, 829–843,
3 doi:10.1080/08941920.2017.1284291.
4 <https://www.tandfonline.com/doi/full/10.1080/08941920.2017.1284291> (Accessed April 12,
5 2019).
- 6 Cattani, D., and S. Asselin, 2018: Has Selection for Grain Yield Altered Intermediate Wheatgrass?
7 *Sustainability*, **10**, 688, doi:10.3390/su10030688. <http://www.mdpi.com/2071-1050/10/3/688>
8 (Accessed May 17, 2018).
- 9 Cavan, G., and Coauthors, 2014: Urban morphological determinants of temperature regulating
10 ecosystem services in two African cities. *Ecol. Indic.*, **42**, 43–57,
11 doi:10.1016/J.ECOLIND.2014.01.025.
12 <https://www.sciencedirect.com/science/article/pii/S1470160X14000338> (Accessed May 17,
13 2018).
- 14 Cayuela, M. L., S. Jeffery, and L. Van Zwieten, 2015: The molar H: Corg ratio of biochar is a key
15 factor in mitigating N₂O emissions from soil. *Agric. Ecosyst. Environ.*, **202**, 135–138.
- 16 Cazenave, A., and G. Le Cozannet, 2014: Sea level rise and its coastal impacts. *Earth's Futur.*, **2**, 15–
17 34, doi:10.1002/2013EF000188. <http://doi.wiley.com/10.1002/2013EF000188> (Accessed May
18 31, 2018).
- 19 CEC, 2009: *Adapting to Climate Change: towards a European Framework for Action*. European
20 Commission, Brussels.
- 21 Cerdà, A., J. Rodrigo-Comino, A. Giménez-Morera, and S. D. Keesstra, 2018: Hydrological and
22 erosional impact and farmer's perception on catch crops and weeds in citrus organic farming in
23 Canyoles river watershed, Eastern Spain. *Agric. Ecosyst. Environ.*, **258**, 49–58,
24 doi:10.1016/J.AGEE.2018.02.015.
25 <https://www.sciencedirect.com/science/article/pii/S0167880918300823> (Accessed April 17,
26 2019).
- 27 Certini, G., 2005: Effects of fire on properties of forest soils: a review. *Oecologia*, **143**, 1–10,
28 doi:10.1007/s00442-004-1788-8. <http://link.springer.com/10.1007/s00442-004-1788-8>
29 (Accessed February 14, 2018).
- 30 Chambers, R., and G. Conway, 1992: *Sustainable rural livelihoods : practical concepts for the 21st*
31 *century*. Institute of Development Studies, 42 pp.
32 <https://opendocs.ids.ac.uk/opendocs/handle/123456789/775> (Accessed January 9, 2018).
- 33 Chan, K. Y., L. Van Zwieten, I. Meszaros, A. Downie, and S. Joseph, 2008: Using poultry litter
34 biochars as soil amendments. *Soil Res.*, **46**, 437, doi:10.1071/SR08036.
35 <http://www.publish.csiro.au/?paper=SR08036> (Accessed April 2, 2019).
- 36 Chandio, N. H., M. M. Anwar, and A. A. Chandio, 2011: Degradation of Indus Delta, Removal of
37 Mangroves Forestland; its causes. A Case Study of Indus River Delta. *Sindh Univ. Res. Jour.*
38 (*Sci. Ser.*), **43**.
- 39 Changjin, S., and C. L., 2007: A review of watershed environmental services from forest in China.
40 Chapter 3. in L. Xiaoyun, J. Leshan, Z. Ting, and I. Bond, editors. *Payment for watershed*
41 *services in China: the role of government and market*. Social Sciences Academic Press (China),
42 Beijing, China.
- 43 Chappell, A., J. Baldock, and J. Sanderman, 2016: The global significance of omitting soil erosion
44 from soil organic carbon cycling schemes. *Nat. Clim. Chang.*, **6**, 187–191,
45 doi:10.1038/nclimate2829. <http://www.nature.com/articles/nclimate2829> (Accessed March 16,
46 2019).
- 47 Charles, R. L., P. K. T. Munushi, and E. F. Nzunda, 2013: Agroforestry as Adaptation Strategy under
48 Climate Change in Mwanza District, Kilimanjaro, Tanzania. *Int. J. Environ. Prot.*, **3**, 29–38.

- 1 Chaudhuri, P., S. Ghosh, M. Bakshi, S. Bhattacharyya, and B. Nath, 2015: A Review of Threats and
2 Vulnerabilities to Mangrove Habitats: With Special Emphasis on East Coast of India. *J. Earth*
3 *Sci. Clim. Change*, **06**, 270, doi:10.4172/2157-7617.1000270. [http://dx.doi.org/10.4172/2157-](http://dx.doi.org/10.4172/2157-7617.1000270)
4 [7617.1000270](http://dx.doi.org/10.4172/2157-7617.1000270) (Accessed July 30, 2018).
- 5 Chazdon, R. L., and M. Uriarte, 2016: Natural regeneration in the context of large-scale forest and
6 landscape restoration in the tropics. *Biotropica*, **48**, 709–715, doi:10.1111/btp.12409.
- 7 Chazdon, R. L., and Coauthors, 2016a: Carbon sequestration potential of second-growth forest
8 regeneration in the Latin American tropics. *Sci. Adv.*, **2**, e1501639–e1501639,
9 doi:10.1126/sciadv.1501639.
- 10 —, and Coauthors, 2016b: Carbon sequestration potential of second-growth forest regeneration in
11 the Latin American tropics. *Sci. Adv.*, **2**, e1501639–e1501639, doi:10.1126/sciadv.1501639.
12 <http://advances.sciencemag.org/cgi/doi/10.1126/sciadv.1501639> (Accessed October 31, 2018).
- 13 Chazdon, R. L., P. H. S. Brancalion, D. Lamb, L. Laestadius, M. Calmon, and C. Kumar, 2017: A
14 Policy-Driven Knowledge Agenda for Global Forest and Landscape Restoration. *Conserv. Lett.*,
15 **10**, 125–132, doi:10.1111/conl.12220. <http://doi.wiley.com/10.1111/conl.12220> (Accessed April
16 11, 2019).
- 17 Cheesman, S., C. Thierfelder, N. S. Eash, G. T. Kassie, and E. Frossard, 2016: Soil carbon stocks in
18 conservation agriculture systems of Southern Africa. *Soil Tillage Res.*, **156**, 99–109,
19 doi:10.1016/J.STILL.2015.09.018.
20 <https://www.sciencedirect.com/science/article/pii/S0167198715300350> (Accessed March 15,
21 2019).
- 22 Chen, C., and Coauthors, 2019: China and India lead in greening of the world through land-use
23 management. *Nat. Sustain.*, **2**, 122–129, doi:10.1038/s41893-019-0220-7.
24 <http://www.nature.com/articles/s41893-019-0220-7> (Accessed February 21, 2019).
- 25 Chen, D., W. Wei, and L. Chen, 2017: Effects of terracing practices on water erosion control in
26 China: A meta-analysis. *Earth-Science Rev.*, **173**, 109–121,
27 doi:10.1016/J.EARSCIREV.2017.08.007.
28 <https://www.sciencedirect.com/science/article/pii/S0012825217300090> (Accessed October 28,
29 2018).
- 30 Chen, J., and Coauthors, 2018a: Prospects for the sustainability of social-ecological systems (SES) on
31 the Mongolian plateau: five critical issues. *Environ. Res. Lett.*, **13**, 123004, doi:10.1088/1748-
32 9326/aaf27b.
- 33 Chen, J., M. T. Ter-Mikaelian, P. Q. Ng, and S. J. Colombo, 2018b: Ontario’s managed forests and
34 harvested wood products contribute to greenhouse gas mitigation from 2020 to 2100. *For.*
35 *Chron.*, **43**, 269–282, doi:10.5558/tfc2018-040.
- 36 Chen, S., and Coauthors, 2018c: Plant diversity enhances productivity and soil carbon storage. *Proc.*
37 *Natl. Acad. Sci. U. S. A.*, **115**, 4027–4032, doi:10.1073/pnas.1700298114.
38 <http://www.ncbi.nlm.nih.gov/pubmed/29666315> (Accessed May 21, 2018).
- 39 Chen, Z., and Coauthors, 2018d: Source Partitioning of Methane Emissions and its Seasonality in the
40 U.S. Midwest. *J. Geophys. Res. Biogeosciences*, **123**, 646–659, doi:10.1002/2017JG004356.
41 <http://doi.wiley.com/10.1002/2017JG004356> (Accessed March 15, 2019).
- 42 Cheng, L., and A. AghaKouchak, 2015: Nonstationary Precipitation Intensity-Duration-Frequency
43 Curves for Infrastructure Design in a Changing Climate. *Sci. Rep.*, **4**, 7093,
44 doi:10.1038/srep07093. <http://www.nature.com/articles/srep07093> (Accessed March 7, 2019).
- 45 Cherlet, M., C. Hutchinson, J. Reynolds, J. Hill, S. Sommer, and G. von Maltitz, 2018: *World Atlas of*
46 *Desertification*. 3rd editio. Publication Office of the European Union, Luxemburg, 248 pp.
- 47 Chia, E., K. Fobissie, M. Kanninen, E. L. Chia, K. Fobissie, and M. Kanninen, 2016: Exploring
48 Opportunities for Promoting Synergies between Climate Change Adaptation and Mitigation in

- 1 Forest Carbon Initiatives. *Forests*, **7**, 24, doi:10.3390/f7010024. <http://www.mdpi.com/1999-4907/7/1/24> (Accessed March 22, 2019).
- 2
- 3 Cho, M.-H., A.-R. Yang, E.-H. Baek, S. M. Kang, S.-J. Jeong, J. Y. Kim, and B.-M. Kim, 2018:
4 Vegetation-cloud feedbacks to future vegetation changes in the Arctic regions. *Clim. Dyn.*, **50**,
5 3745–3755, doi:10.1007/s00382-017-3840-5. <http://link.springer.com/10.1007/s00382-017-3840-5> (Accessed March 16, 2019).
- 6
- 7 Choi, S.-D., K. Lee, and Y.-S. Chang, 2002: Large rate of uptake of atmospheric carbon dioxide by
8 planted forest biomass in Korea. *Global Biogeochem. Cycles*, **16**, 1089,
9 doi:10.1029/2002GB001914, doi:10.1029/2002GB001914.
10 <http://doi.wiley.com/10.1029/2002GB001914> (Accessed May 21, 2018).
- 11 Christensen, T. R., T. Johansson, H. J. Åkerman, M. Mastepanov, N. Malmer, T. Friborg, P. Crill, and
12 B. H. Svensson, 2004: Thawing sub-arctic permafrost: Effects on vegetation and methane
13 emissions. *Geophys. Res. Lett.*, **31**, L04501, doi:10.1029/2003GL018680.
14 <http://doi.wiley.com/10.1029/2003GL018680> (Accessed March 16, 2019).
- 15 Ciais, P., and Coauthors, 2013: Carbon and Other Biogeochemical Cycles. *Climate Change 2013: The
16 Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the
17 Intergovernmental Panel on Climate Change*, J. Stocker, T.F., D. Qin, G.-K. Plattner, M.
18 Tignor, S.K. Allen and V.B. and P.M.M. Boschung, A. Nauels, Y. Xia, Eds., Cambridge
19 University Press, Cambridge, UK and New York, USA, 467–570.
- 20 Cilas, C., F.-R. Goebel, R. Babin, and J. Avelino, 2016: Tropical Crop Pests and Diseases in a
21 Climate Change Setting—A Few Examples. *Climate Change and Agriculture Worldwide*,
22 Springer Netherlands, Dordrecht, 73–82 http://link.springer.com/10.1007/978-94-017-7462-8_6
23 (Accessed March 18, 2019).
- 24 Classen, A. T., M. K. Sundqvist, J. A. Henning, G. S. Newman, J. A. M. Moore, M. A. Cregger, L. C.
25 Moorhead, and C. M. Patterson, 2015: Direct and indirect effects of climate change on soil
26 microbial and soil microbial-plant interactions: What lies ahead? *Ecosphere*, **6**, art130,
27 doi:10.1890/ES15-00217.1. <http://doi.wiley.com/10.1890/ES15-00217.1> (Accessed May 2,
28 2018).
- 29 Cohen-Shacham, E., Walters, G., Janzen, C. and Maginnis, S., 2016: *Nature-based Solutions to
30 address global societal challenges*. Gland, Switzerland, xiii + 97pp. pp.
- 31 Colloff, M. J., S. Lavorel, R. M. Wise, M. Dunlop, I. C. Overton, and K. J. Williams, 2016:
32 Adaptation services of floodplains and wetlands under transformational climate change. *Ecol.
33 Appl.*, **26**, 1003–1017, doi:10.1890/15-0848. <http://doi.wiley.com/10.1890/15-0848> (Accessed
34 April 16, 2019).
- 35 Colombani, N., A. Osti, G. Volta, and M. Mastrocicco, 2016: Impact of Climate Change on
36 Salinization of Coastal Water Resources. *Water Resour. Manag.*, **30**, 2483–2496,
37 doi:10.1007/s11269-016-1292-z. <http://link.springer.com/10.1007/s11269-016-1292-z>
38 (Accessed February 28, 2019).
- 39 Coma, J., G. Pírez, A. de Gracia, S. Burić, M. Urrestarazu, and L. F. Cabeza, 2017: Vertical
40 greenery systems for energy savings in buildings: A comparative study between green walls and
41 green facades. *Build. Environ.*, doi:10.1016/j.buildenv.2016.11.014.
- 42 Conant, R. T., and K. Paustian, 2002: Potential soil carbon sequestration in overgrazed grassland
43 ecosystems. *Global Biogeochem. Cycles*, **16**, 90-1-90–99, doi:10.1029/2001GB001661.
44 <http://doi.wiley.com/10.1029/2001GB001661> (Accessed October 3, 2018).
- 45 —, and Coauthors, 2011a: Temperature and soil organic matter decomposition rates - synthesis of
46 current knowledge and a way forward. *Glob. Chang. Biol.*, **17**, 3392–3404, doi:10.1111/j.1365-
47 2486.2011.02496.x. <http://doi.wiley.com/10.1111/j.1365-2486.2011.02496.x> (Accessed
48 February 14, 2018).
- 49 Conant, R. T., S. M. Ogle, E. A. Paul, and K. Paustian, 2011b: Measuring and monitoring soil organic

- 1 carbon stocks in agricultural lands for climate mitigation. *Front. Ecol. Environ.*, **9**, 169–173,
2 doi:10.1890/090153. <http://doi.wiley.com/10.1890/090153> (Accessed May 23, 2018).
- 3 Conant, R. T., C. E. P. Cerri, B. B. Osborne, and K. Paustian, 2017: Grassland management impacts
4 on soil carbon stocks: a new synthesis. *Ecol. Appl.*, **27**, 662–668, doi:10.1002/eap.1473.
5 <http://doi.wiley.com/10.1002/eap.1473> (Accessed April 12, 2019).
- 6 Coomes, O. T., Y. Takasaki, and J. M. Rhemtulla, 2011: Land-use poverty traps identified in shifting
7 cultivation systems shape long-term tropical forest cover. *Proc. Natl. Acad. Sci.*, **108**, 13925–
8 13930, doi:10.1073/PNAS.1012973108. <https://www.pnas.org/content/108/34/13925.long>
9 (Accessed March 5, 2019).
- 10 Cooper, J. A. G., and J. Pile, 2014: The adaptation-resistance spectrum: A classification of
11 contemporary adaptation approaches to climate-related coastal change. *Ocean Coast. Manag.*,
12 **94**, 90–98, doi:10.1016/j.ocecoaman.2013.09.006.
- 13 Coppus, R., and A. C. Imeson, 2002: Extreme events controlling erosion and sediment transport in a
14 semi-arid sub-andean valley. *Earth Surf. Process. Landforms*, **27**, 1365–1375,
15 doi:10.1002/esp.435. <http://doi.wiley.com/10.1002/esp.435> (Accessed May 22, 2018).
- 16 Costanza, R., O. Pérez-Maqueo, M. L. Martinez, P. Sutton, S. J. Anderson, and K. Mulder, 2008: The
17 Value of Coastal Wetlands for Hurricane Protection. *AMBIO A J. Hum. Environ.*, **37**, 241–248,
18 doi:10.1579/0044-7447(2008)37[241:TVOCWF]2.0.CO;2.
19 [http://www.bioone.org/doi/abs/10.1579/0044-7447\(2008\)37\[241:TVOCWF\]2.0.CO;2](http://www.bioone.org/doi/abs/10.1579/0044-7447(2008)37[241:TVOCWF]2.0.CO;2) (Accessed
20 May 24, 2018).
- 21 Cote, M., and A. J. Nightingale, 2012: Resilience thinking meets social theory. *Prog. Hum. Geogr.*,
22 **36**, 475–489, doi:10.1177/0309132511425708.
23 <http://journals.sagepub.com/doi/10.1177/0309132511425708> (Accessed May 25, 2018).
- 24 Cotrufo, M. F., J. L. Soong, A. J. Horton, E. E. Campbell, M. L. Haddix, D. H. Wall, and W. J.
25 Parton, 2015: Formation of soil organic matter via biochemical and physical pathways of litter
26 mass loss. *Nat. Geosci.*, **8**, 776–779, doi:10.1038/ngeo2520.
27 <http://www.nature.com/articles/ngeo2520> (Accessed March 25, 2019).
- 28 Cowie, A., A. D. Woolf, J. Gaunt, M. Brandão, R. A. de la Rosa, and C. A., 2015: Biochar, carbon
29 accounting and climate change. *Biochar for Environmental Management Science, Technology*
30 *and Implementation Edited By Stephen Joseph, Johannes Lehmann. Taylor and Francis. 763-*
31 *794.*
- 32 Cowie, A. L., and Coauthors, 2018: Land in balance: The scientific conceptual framework for Land
33 Degradation Neutrality. *Environ. Sci. Policy*, **79**, 25–35.
- 34 Cox, S., P. Nabukalu, A. Paterson, W. Kong, and S. Nakasagga, 2018: Development of Perennial
35 Grain Sorghum. *Sustainability*, **10**, 172, doi:10.3390/su10010172. [http://www.mdpi.com/2071-](http://www.mdpi.com/2071-1050/10/1/172)
36 [1050/10/1/172](http://www.mdpi.com/2071-1050/10/1/172) (Accessed May 21, 2018).
- 37 Cramer, W., and Coauthors, 2014: Detection and attribution of observed impacts. *Climate Change*
38 *2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects.*
39 *Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental*
40 *Panel on Climate Change*, D.J.D. [Field, C.B., V.R. Barros, E.S.K. K.J. Mach, M.D.
41 Mastrandrea, T.E. Bilir, M. Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova, B. Girma, and
42 And L.L.W. A.N. Levy, S. MacCracken, P.R. Mastrandrea, Eds., Cambridge University Press,
43 Cambridge, UK and New York, USA, 979–1037.
- 44 Cretney, R., 2014: Resilience for Whom? Emerging Critical Geographies of Socio-ecological
45 Resilience. *Geogr. Compass*, **8**, 627–640, doi:10.1111/gec3.12154.
46 <http://doi.wiley.com/10.1111/gec3.12154> (Accessed May 25, 2018).
- 47 Crews, T., W. Carton, and L. Olsson, 2018: Is the future of agriculture perennial? Imperatives and
48 opportunities to reinvent agriculture by shifting from annual monocultures to perennial
49 polycultures. *Glob. Sustain.*, **forthcomin**, 35.

- 1 Crews, T. E., and S. R. Gliessman, 1991: Raised field agriculture in Tlaxcala, Mexico: An ecosystem
2 perspective on maintenance of soil fertility. *Am. J. Altern. Agric.*, **6**, 9,
3 doi:10.1017/S088918930000374X.
4 http://www.journals.cambridge.org/abstract_S088918930000374X (Accessed October 29,
5 2018).
- 6 ———, and B. E. Rumsey, 2017: What agriculture can learn from native ecosystems in building soil
7 organic matter: A review. *Sustain.*, **9**, 1–18, doi:10.3390/su9040578.
- 8 Crews, T. E., J. Blesh, S. W. Culman, R. C. Hayes, E. Steen Jensen, M. C. Mack, M. B. Peoples, and
9 M. E. Schipanski, 2016: Going where no grains have gone before: From early to mid-succession.
10 *Agric. Ecosyst. Environ.*, **223**, 223–238, doi:10.1016/j.agee.2016.03.012.
11 <http://dx.doi.org/10.1016/j.agee.2016.03.012> (Accessed October 31, 2017).
- 12 Crowther, T. W., S. M. Thomas, D. S. Maynard, P. Baldrian, K. Covey, S. D. Frey, L. T. A. van
13 Diepen, and M. A. Bradford, 2015: Biotic interactions mediate soil microbial feedbacks to
14 climate change. *Proc. Natl. Acad. Sci.*, **112**, 7033–7038, doi:10.1073/pnas.1502956112.
15 <http://www.pnas.org/lookup/doi/10.1073/pnas.1502956112> (Accessed October 1, 2018).
- 16 Crowther, T. W., and Coauthors, 2016: Quantifying global soil carbon losses in response to warming.
17 *Nature*, **540**, 104–108, doi:10.1038/nature20150. <http://www.nature.com/articles/nature20150>
18 (Accessed February 27, 2019).
- 19 Crozier, M. J., 2010: Deciphering the effect of climate change on landslide activity: A review.
20 *Geomorphology*, **124**, 260–267, doi:10.1016/J.GEOMORPH.2010.04.009.
21 <https://www.sciencedirect.com/science/article/pii/S0169555X10001881> (Accessed April 10,
22 2019).
- 23 Čuček, L., J. J. Klemeš, and Z. Kravanja, 2012: A Review of Footprint analysis tools for monitoring
24 impacts on sustainability. *J. Clean. Prod.*, **34**, 9–20, doi:10.1016/J.JCLEPRO.2012.02.036.
25 <https://www.sciencedirect.com/science/article/pii/S0959652612001126> (Accessed October 25,
26 2018).
- 27 Cui, G., and Coauthors, 2014: Estimation of forest carbon budget from land cover change in South
28 and North Korea between 1981 and 2010. *J. Plant Biol.*, **57**, 225–238, doi:10.1007/s12374-014-
29 0165-3. <http://link.springer.com/10.1007/s12374-014-0165-3> (Accessed May 21, 2018).
- 30 Culman, S. W., S. S. Snapp, M. Ollenburger, B. Basso, and L. R. DeHaan, 2013: Soil and water
31 quality rapidly responds to the perennial grain Kernza wheatgrass. *Agron. J.*, **105**, 735–744,
32 doi:10.2134/agronj2012.0273.
- 33 Curtis, P. G., C. M. Slay, N. L. Harris, A. Tyukavina, and M. C. Hansen, 2018: Classifying drivers of
34 global forest loss. *Science (80-.)*, **361**, 1108–1111, doi:10.1126/science.aau3445.
35 <http://www.sciencemag.org/lookup/doi/10.1126/science.aau3445>.
- 36 Cutter, S. L., L. Barnes, M. Berry, C. Burton, E. Evans, E. Tate, and J. Webb, 2008: A place-based
37 model for understanding community resilience to natural disasters. *Glob. Environ. Chang.*, **18**,
38 598–606, doi:10.1016/J.GLOENVCHA.2008.07.013.
39 <https://www.sciencedirect.com/science/article/pii/S0959378008000666> (Accessed May 25,
40 2018).
- 41 D’Annunzio, R., E. J. Lindquist, and K. G. MacDicken, 2017: *Global forest land- use change from*
42 *1990 to 2010: an update to a global remote sensing survey of forests*. FAO, Rome, Italy,
43 <http://www.fao.org/3/a-i5098e.pdf>.
- 44 Dallimer, M., and L. C. Stringer, 2018: Informing investments in land degradation neutrality efforts:
45 A triage approach to decision making. *Environ. Sci. Policy*, **89**, 198–205,
46 doi:10.1016/j.envsci.2018.08.004.
- 47 Dargie, G. C., S. L. Lewis, I. T. Lawson, E. T. A. Mitchard, S. E. Page, Y. E. Bocko, and S. A. Ifo,
48 2017: Age, extent and carbon storage of the central Congo Basin peatland complex. *Nature*, **542**,
49 86–90, doi:10.1038/nature21048.

- 1 Darrah, S. E., Y. Shennan-Farpón, J. Loh, N. C. Davidson, C. M. Finlayson, R. C. Gardner, and M. J.
2 Walpole, 2019: Improvements to the Wetland Extent Trends (WET) index as a tool for
3 monitoring natural and human-made wetlands. *Ecol. Indic.*, **99**, 294–298,
4 doi:10.1016/J.ECOLIND.2018.12.032.
5 <https://www.sciencedirect.com/science/article/pii/S1470160X18309671> (Accessed April 16,
6 2019).
- 7 Davidson, N. C., 2014: How much wetland has the world lost? Long-term and recent trends in global
8 wetland area. *Mar. Freshw. Res.*, **65**, 934, doi:10.1071/MF14173.
9 <http://www.publish.csiro.au/?paper=MF14173> (Accessed April 16, 2019).
- 10 Davies, M. I. J., 2015: Economic Specialisation, Resource Variability, and the Origins of Intensive
11 Agriculture in Eastern Africa. *Rural Landscapes Soc. Environ. Hist.*, **2**, doi:10.16993/rl.af.
12 <http://www.rurallandscapesjournal.com/article/view/rl.af/> (Accessed March 27, 2019).
- 13 —, and H. L. Moore, 2016: Landscape, time and cultural resilience: a brief history of agriculture in
14 Pokot and Marakwet, Kenya. *J. East. African Stud.*, **10**, 67–87,
15 doi:10.1080/17531055.2015.1134417.
16 <http://www.tandfonline.com/doi/full/10.1080/17531055.2015.1134417> (Accessed March 27,
17 2019).
- 18 Davies, Z. G., J. L. Edmondson, A. Heinemeyer, J. R. Leake, and K. J. Gaston, 2011: Mapping an
19 urban ecosystem service: Quantifying above-ground carbon storage at a city-wide scale. *J. Appl.*
20 *Ecol.*, doi:10.1111/j.1365-2664.2011.02021.x.
- 21 Davin, E. L., N. de Noblet-Ducoudré, E. L. Davin, and N. de Noblet-Ducoudré, 2010: Climatic
22 Impact of Global-Scale Deforestation: Radiative versus Nonradiative Processes. *J. Clim.*, **23**,
23 97–112, doi:10.1175/2009JCLI3102.1.
24 <http://journals.ametsoc.org/doi/abs/10.1175/2009JCLI3102.1> (Accessed May 12, 2018).
- 25 Davin, E. L., S. I. Seneviratne, P. Ciais, A. Olioso, and T. Wang, 2014: Preferential cooling of hot
26 extremes from cropland albedo management. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 9757–9761,
27 doi:10.1073/pnas.1317323111. <http://www.ncbi.nlm.nih.gov/pubmed/24958872> (Accessed May
28 12, 2018).
- 29 Day, J. J., and K. I. Hodges, 2018: Growing Land-Sea Temperature Contrast and the Intensification of
30 Arctic Cyclones. *Geophys. Res. Lett.*, **45**, 3673–3681, doi:10.1029/2018GL077587.
31 <http://doi.wiley.com/10.1029/2018GL077587> (Accessed March 14, 2019).
- 32 DeConto, R. M., and D. Pollard, 2016: Contribution of Antarctica to past and future sea-level rise.
33 *Nature*, **531**, 591–597, doi:10.1038/nature17145. <http://www.nature.com/articles/nature17145>
34 (Accessed March 14, 2019).
- 35 DeHaan, L., M. Christians, J. Crain, and J. Poland, 2018: Development and Evolution of an
36 Intermediate Wheatgrass Domestication Program. *Sustainability*, **10**, 1499,
37 doi:10.3390/su10051499. <http://www.mdpi.com/2071-1050/10/5/1499> (Accessed May 21,
38 2018).
- 39 DeHaan, L. R., and D. L. Van Tassel, 2014: Useful insights from evolutionary biology for developing
40 perennial grain crops1. *Am. J. Bot.*, **101**, 1801–1819, doi:10.3732/ajb.1400084.
41 <http://doi.wiley.com/10.3732/ajb.1400084> (Accessed May 21, 2018).
- 42 DeHaan, L. R., and Coauthors, 2016: A pipeline strategy for grain crop domestication. *Crop Sci.*, **56**,
43 917–930, doi:10.2135/cropsci2015.06.0356.
- 44 Delgado, A., and J. A. Gómez, 2016: The Soil. Physical, Chemical and Biological Properties.
45 *Principles of Agronomy for Sustainable Agriculture*, Springer International Publishing, Cham,
46 15–26 http://link.springer.com/10.1007/978-3-319-46116-8_2 (Accessed March 4, 2019).
- 47 Dell, M., B. F. Jones, and B. A. Olken, 2009: Temperature and Income: Reconciling New Cross-
48 Sectional and Panel Estimates. *Am. Econ. Rev.*, **99**, 198–204, doi:10.1257/aer.99.2.198.
49 <http://pubs.aeaweb.org/doi/10.1257/aer.99.2.198> (Accessed March 8, 2019).

- 1 Demuzere, M., and Coauthors, 2014: Mitigating and adapting to climate change: Multi-functional and
2 multi-scale assessment of green urban infrastructure. *J. Environ. Manage.*, **146**, 107–115,
3 doi:10.1016/J.JENVMAN.2014.07.025.
4 <https://www.sciencedirect.com/science/article/pii/S0301479714003740> (Accessed May 17,
5 2018).
- 6 Devendra, C., 2002: Crop–animal systems in Asia: implications for research. *Agric. Syst.*, **71**, 169–
7 177, doi:10.1016/S0308-521X(01)00042-7.
8 <https://www.sciencedirect.com/science/article/pii/S0308521X01000427> (Accessed April 2,
9 2019).
- 10 —, and D. Thomas, 2002: Crop–animal systems in Asia: importance of livestock and
11 characterisation of agro-ecological zones. *Agric. Syst.*, **71**, 5–15, doi:10.1016/S0308-
12 521X(01)00032-4. <https://www.sciencedirect.com/science/article/pii/S0308521X01000324>
13 (Accessed April 2, 2019).
- 14 DG Environment News Alert Service, 2012: *The Multifunctionality of Green Infrastructure*. European
15 Commission, Brussels, Belgium, 40 pp.
16 http://ec.europa.eu/environment/nature/ecosystems/docs/Green_Infrastructure.pdf (Accessed
17 May 17, 2018).
- 18 Dhanush, D., and Coauthors, 2015: Impact of climate change on African agriculture: focus on pests
19 and diseases. <https://cgspace.cgiar.org/handle/10568/66472> (Accessed October 24, 2018).
- 20 Dignac, M.-F., and Coauthors, 2017: Increasing soil carbon storage: mechanisms, effects of
21 agricultural practices and proxies. A review. *Agron. Sustain. Dev.*, **37**, 14, doi:10.1007/s13593-
22 017-0421-2. <http://link.springer.com/10.1007/s13593-017-0421-2> (Accessed May 11, 2018).
- 23 Ding, F., L. Van Zwieten, W. Zhang, Z. (Han) Weng, S. Shi, J. Wang, and J. Meng, 2018: A meta-
24 analysis and critical evaluation of influencing factors on soil carbon priming following biochar
25 amendment. *J. Soils Sediments*, doi:10.1007/s11368-017-1899-6.
- 26 Dintwe, K., G. S. Okin, and Y. Xue, 2017: Fire-induced albedo change and surface radiative forcing
27 in sub-Saharan Africa savanna ecosystems: Implications for the energy balance. *J. Geophys.*
28 *Res. Atmos.*, **122**, 6186–6201, doi:10.1002/2016JD026318.
29 <http://doi.wiley.com/10.1002/2016JD026318> (Accessed May 12, 2018).
- 30 Dixon, M. J. R., J. Loh, N. C. Davidson, C. Beltrame, R. Freeman, and M. Walpole, 2016: Tracking
31 global change in ecosystem area: The Wetland Extent Trends index. *Biol. Conserv.*, **193**, 27–35,
32 doi:10.1016/j.biocon.2015.10.023.
33 <https://linkinghub.elsevier.com/retrieve/pii/S0006320715301476> (Accessed April 16, 2019).
- 34 Djurfeldt, A. A., E. Hillbom, W. O. Mulwafu, P. Mvula, and G. Djurfeldt, 2018: “The family farms
35 together, the decisions, however are made by the man” —Matrilineal land tenure systems,
36 welfare and decision making in rural Malawi. *Land use policy*, **70**, 601–610,
37 doi:10.1016/j.landusepol.2017.10.048.
38 <https://linkinghub.elsevier.com/retrieve/pii/S0264837717306683> (Accessed October 18, 2018).
- 39 Dommain, R., A. Barthelmes, F. Tanneberger, A. Bonn, C. Bain, and H. Joosten, 2012: Peatlands -
40 guidance for climate change mitigation through conservation, rehabilitation and sustainable use.
41 *Mitigation of Climate Change in Agriculture (MICCA) Programme series 5*.
- 42 —, S. Frolking, A. Jeltsch-Thömmes, F. Joos, J. Couwenberg, and P. H. Glaser, 2018: A radiative
43 forcing analysis of tropical peatlands before and after their conversion to agricultural
44 plantations. *Glob. Chang. Biol.*, **24**, 5518–5533, doi:10.1111/gcb.14400.
- 45 Donner, S., 2012: Sea level rise and the ongoing Battle of Tarawa. *Eos, Trans. Am. Geophys. Union*,
46 **93**, 169–170, doi:10.1029/2012EO170001.
- 47 Donner, S. D., and S. Webber, 2014: Obstacles to climate change adaptation decisions: a case study of
48 sea-level rise and coastal protection measures in Kiribati. *Sustain. Sci.*, **9**, 331–345,
49 doi:10.1007/s11625-014-0242-z.

- 1 Dooley, K., and S. Kartha, 2018: Land-based negative emissions: risks for climate mitigation and
2 impacts on sustainable development. *Int. Environ. Agreements Polit. Law Econ.*, **18**, 79–98,
3 doi:10.1007/s10784-017-9382-9. <http://link.springer.com/10.1007/s10784-017-9382-9>
4 (Accessed March 16, 2018).
- 5 Doss, C., C. Kovarik, A. Peterman, A. Quisumbing, and M. van den Bold, 2015: Gender inequalities
6 in ownership and control of land in Africa: myth and reality. *Agric. Econ.*, **46**, 403–434,
7 doi:10.1111/agec.12171. <http://doi.wiley.com/10.1111/agec.12171> (Accessed October 19, 2018).
- 8 Dotterweich, M., 2008: The history of soil erosion and fluvial deposits in small catchments of central
9 Europe: Deciphering the long-term interaction between humans and the environment — A
10 review. *Geomorphology*, **101**, 192–208, doi:10.1016/J.GEOMORPH.2008.05.023.
11 <https://www.sciencedirect.com/science/article/pii/S0169555X08002213#bib100> (Accessed
12 February 27, 2019).
- 13 —, 2013: The history of human-induced soil erosion: Geomorphic legacies, early descriptions and
14 research, and the development of soil conservation—A global synopsis. *Geomorphology*, **201**,
15 1–34, doi:10.1016/J.GEOMORPH.2013.07.021.
16 <https://www.sciencedirect.com/science/article/pii/S0169555X1300370X> (Accessed February 27,
17 2019).
- 18 Dow, K., F. Berkhout, and B. L. Preston, 2013a: Limits to adaptation to climate change: a risk
19 approach. *Curr. Opin. Environ. Sustain.*, **5**, 384–391, doi:10.1016/J.COSUST.2013.07.005.
20 <https://www.sciencedirect.com/science/article/pii/S1877343513000845> (Accessed March 29,
21 2019).
- 22 —, —, B. L. Preston, R. J. T. Klein, G. Midgley, and M. R. Shaw, 2013b: Limits to adaptation.
23 *Nat. Clim. Chang.*, **3**, 305–307, doi:10.1038/nclimate1847.
24 <http://www.nature.com/articles/nclimate1847> (Accessed November 2, 2018).
- 25 Downie, A., P. Munroe, A. Cowie, L. Van Zwieten, and D. M. S. Lau, 2012: Biochar as a
26 Geoenvironmental Climate Solution: Hazard Identification and Risk Management. *Crit. Rev.*
27 *Environ. Sci. Technol.*, doi:10.1080/10643389.2010.507980.
- 28 Dragoni, D., H. P. Schmid, C. A. Wayson, H. Potters, C. S. B. Grimmond, and J. C. Randolph, 2011:
29 Evidence of increased net ecosystem productivity associated with a longer vegetated season in a
30 deciduous forest in south-central Indiana, USA. *Glob. Chang. Biol.*, **17**, 886–897,
31 doi:10.1111/j.1365-2486.2010.02281.x.
- 32 Dregne, H. E., 1998: Desertification Assessment. *Methods for assessment of soil degradation*, R. Lal,
33 Ed., CRC Press, Boca Raton, London, New York, Washington DC, 441–458.
- 34 Drescher, J., and Coauthors, 2016: Ecological and socio-economic functions across tropical land use
35 systems after rainforest conversion. *Philos. Trans. R. Soc. B Biol. Sci.*, **371**, 20150275,
36 doi:10.1098/rstb.2015.0275.
37 <http://rstb.royalsocietypublishing.org/lookup/doi/10.1098/rstb.2015.0275> (Accessed March 14,
38 2019).
- 39 Drösler, M., L. V. Verchot, A. Freibauer, and G. Pan, 2013: Drained inland organic soils. *2013 Suppl.*
40 *to 2006 IPCC Guidel. Natl. Greenh. Gas Invent. Wetl.*, 1–79.
- 41 Dugan, A. J., R. Birdsey, V. S. Mascorro, M. Magnan, C. E. Smyth, M. Olguin, and W. A. Kurz,
42 2018: A systems approach to assess climate change mitigation options in landscapes of the
43 United States forest sector. *Carbon Balance Manag.*, **13**, 13, doi:10.1186/s13021-018-0100-x.
44 <https://doi.org/10.1186/s13021-018-0100-x>.
- 45 Dumanski, J., and C. Pieri, 2000: Land quality indicators: research plan. *Agric. Ecosyst. Environ.*, **81**,
46 93–102, doi:10.1016/S0167-8809(00)00183-3.
47 <https://www.sciencedirect.com/science/article/pii/S0167880900001833> (Accessed September
48 27, 2018).
- 49 Duncan, T., 2016: Case Study: Taranaki Farm Regenerative Agriculture. Pathways to Integrated

- 1 Ecological Farming. *L. Restor.*, 271–287, doi:10.1016/B978-0-12-801231-4.00022-7.
2 <https://www.sciencedirect.com/science/article/pii/B9780128012314000227> (Accessed April 14,
3 2019).
- 4 Dupouey, J. L., E. Dambrine, J. D. Laffite, and C. Moares, 2002: Irreversible impact of past land use
5 on forest soils and biodiversity. *Ecology*, **83**, 2978–2984, doi:10.1890/0012-
6 9658(2002)083[2978:IIOPLU]2.0.CO;2.
7 [https://esajournals.onlinelibrary.wiley.com/doi/full/10.1890/0012-
8 9658%282002%29083%5B2978%3AIIOPLU%5D2.0.CO%3B2](https://esajournals.onlinelibrary.wiley.com/doi/full/10.1890/0012-9658%282002%29083%5B2978%3AIIOPLU%5D2.0.CO%3B2) (Accessed May 23, 2018).
- 9 Dutta, D., P. K. Das, S. Paul, J. R. Sharma, and V. K. Dadhwal, 2015: Assessment of ecological
10 disturbance in the mangrove forest of Sundarbans caused by cyclones using MODIS time-series
11 data (2001–2011). *Nat. Hazards*, **79**, 775–790, doi:10.1007/s11069-015-1872-x.
12 <http://link.springer.com/10.1007/s11069-015-1872-x> (Accessed June 19, 2018).
- 13 Duvat, V., 2013: Coastal protection structures in Tarawa Atoll, Republic of Kiribati. *Sustain. Sci.*, **8**,
14 363–379, doi:10.1007/s11625-013-0205-9.
- 15 Duvat, V. K. E., A. K. Magnan, S. Etienne, C. Salmon, and C. Pignon-Mussaud, 2016: Assessing the
16 impacts of and resilience to Tropical Cyclone Bejisa, Reunion Island (Indian Ocean). *Nat.*
17 *Hazards*, **83**, 601–640, doi:10.1007/s11069-016-2338-5.
- 18 Earles, J. M., S. Yeh, and K. E. Skog, 2012: Timing of carbon emissions from global forest clearance.
19 *Nat. Clim. Chang.*, **2**, 682–685, doi:10.1038/nclimate1535.
- 20 Edmondson, J. L., Z. G. Davies, S. A. McCormack, K. J. Gaston, and J. R. Leake, 2011: Are soils in
21 urban ecosystems compacted? A citywide analysis. *Biol. Lett.*, doi:10.1098/rsbl.2011.0260.
- 22 —, —, —, —, and —, 2014: Land-cover effects on soil organic carbon stocks in a
23 European city. *Sci. Total Environ.*, doi:10.1016/j.scitotenv.2013.11.025.
- 24 Edstedt, K., and W. Carton, 2018a: The benefits that (only) capital can see? Resource access and
25 degradation in industrial carbon forestry, lessons from the CDM in Uganda. *Geoforum*, **97**, 315–
26 323, doi:10.1016/J.GEOFORUM.2018.09.030.
27 <https://www.sciencedirect.com/science/article/pii/S0016718518302896> (Accessed April 12,
28 2019).
- 29 —, and —, 2018b: The benefits that (only) capital can see? Resource access and degradation in
30 industrial carbon forestry, lessons from the CDM in Uganda. *Geoforum*,
31 doi:10.1016/J.GEOFORUM.2018.09.030.
32 <https://www.sciencedirect.com/science/article/pii/S0016718518302896> (Accessed November 1,
33 2018).
- 34 Eglin, T., and Coauthors, 2010: Historical and future perspectives of global soil carbon response to
35 climate and land-use changes. *Tellus B Chem. Phys. Meteorol.*, **62**, 700–718,
36 doi:10.1111/j.1600-0889.2010.00499.x. [https://www.tandfonline.com/doi/full/10.1111/j.1600-
37 0889.2010.00499.x](https://www.tandfonline.com/doi/full/10.1111/j.1600-0889.2010.00499.x) (Accessed March 18, 2019).
- 38 Ehnröström, B., 2001: Leaving Dead Wood for Insects in Boreal Forests - Suggestions for the Future.
39 *Scand. J. For. Res.*, **16**, 91–98, doi:10.1080/028275801300090681.
40 <http://www.tandfonline.com/doi/abs/10.1080/028275801300090681> (Accessed May 26, 2018).
- 41 Ehrenfeld, J. G., B. Ravit, and K. Elgersma, 2005: FEEDBACK IN THE PLANT-SOIL SYSTEM.
42 *Annu. Rev. Environ. Resour.*, **30**, 75–115, doi:10.1146/annurev.energy.30.050504.144212.
43 <http://www.annualreviews.org/doi/10.1146/annurev.energy.30.050504.144212> (Accessed
44 October 1, 2018).
- 45 Eisenbies, M. H., W. M. Aust, J. A. Burger, and M. B. Adams, 2007: Forest operations, extreme
46 flooding events, and considerations for hydrologic modeling in the Appalachians—A review.
47 *For. Ecol. Manage.*, **242**, 77–98, doi:10.1016/j.foreco.2007.01.051. [https://ac.els-
48 cdn.com/S0378112707000746/1-s2.0-S0378112707000746-main.pdf?_tid=2ecd3669-3c0e-
49 4f49-b501-b964be8ffd62&acdnat=1520241780_5f53e0bc68da2f0d2220c0d9136aa38b](https://ac.els-cdn.com/S0378112707000746/1-s2.0-S0378112707000746-main.pdf?_tid=2ecd3669-3c0e-4f49-b501-b964be8ffd62&acdnat=1520241780_5f53e0bc68da2f0d2220c0d9136aa38b)

- 1 (Accessed March 5, 2018).
- 2 Elad, Y., E. Cytryn, Y. Meller Harel, B. Lew, and E. R. Graber, 2011: The biochar effect: Plant
3 resistance to biotic stresses. *Phytopathol. Mediterr.*, doi:10.14601/Phytopathol_Mediterr-9807.
- 4 Ellbehri et al., 2017: *FAO-IPCC Expert Meeting on Climate Change, Land Use and Food Security:*
5 *Final Meeting Report. Rome, 156 pp.*
- 6 Elliott, M., N. D. Cutts, and A. Trono, 2014: A typology of marine and estuarine hazards and risks as
7 vectors of change: A review for vulnerable coasts and their management. *Ocean Coast. Manag.*,
8 **93**, 88–99, doi:10.1016/J.OCECOAMAN.2014.03.014.
9 <https://www.sciencedirect.com/science/article/pii/S096456911400074X> (Accessed October 27,
10 2018).
- 11 Ellis, E. C., J. O. Kaplan, D. Q. Fuller, S. Vavrus, K. Klein Goldewijk, and P. H. Verburg, 2013: Used
12 planet: a global history. *Proc. Natl. Acad. Sci. U. S. A.*, **110**, 7978–7985,
13 doi:10.1073/pnas.1217241110. <http://www.ncbi.nlm.nih.gov/pubmed/23630271> (Accessed
14 February 27, 2019).
- 15 Ellis, J. E., 1994: Climate variability and complex ecosystem dynamics: implications for pastoral
16 developmen. *Living with uncertainty: New directions in pastoral development in Africa*, I.
17 Scoones, Ed., Intermediate Technology Publications, London, 37–46.
- 18 Ellis, P. W., and Coauthors, 2019: Reduced-impact logging for climate change mitigation (RIL-C) can
19 halve selective logging emissions from tropical forests. *For. Ecol. Manage.*, **438**, 255–266,
20 doi:10.1016/J.FORECO.2019.02.004.
21 https://www.sciencedirect.com/science/article/pii/S0378112718322126?dgcid=rss_sd_all.
- 22 Ellison, D., and Coauthors, 2017: Trees, forests and water: Cool insights for a hot world. *Glob.*
23 *Environ. Chang.*, **43**, 51–61, doi:10.1016/J.GLOENVCHA.2017.01.002.
24 <https://www.sciencedirect.com/science/article/pii/S0959378017300134> (Accessed March 13,
25 2019).
- 26 Elmhirst, R., 2011: Introducing new feminist political ecologies. *Geoforum*, **42**, 129–132,
27 doi:10.1016/j.geoforum.2011.01.006.
28 <http://linkinghub.elsevier.com/retrieve/pii/S001671851100008X> (Accessed April 16, 2018).
- 29 England, M. H., and Coauthors, 2014: Recent intensification of wind-driven circulation in the Pacific
30 and the ongoing warming hiatus. *Nat. Clim. Chang.*, **4**, 222–227, doi:10.1038/nclimate2106.
31 <http://www.nature.com/articles/nclimate2106> (Accessed March 8, 2019).
- 32 Erb, K.-H. H., and Coauthors, 2018: Unexpectedly large impact of forest management and grazing on
33 global vegetation biomass. *Nature*, **553**, 73–76, doi:10.1038/nature25138.
34 <http://dx.doi.org/10.1038/nature25138> (Accessed May 26, 2018).
- 35 Erwin, K. L., 2009: Wetlands and global climate change: The role of wetland restoration in a
36 changing world. *Wetl. Ecol. Manag.*, **17**, 71–84, doi:10.1007/s11273-008-9119-1.
37 [https://www.wetlands.org/wp-content/uploads/2015/11/Wetlands-and-Global-Climate-](https://www.wetlands.org/wp-content/uploads/2015/11/Wetlands-and-Global-Climate-Change.pdf)
38 [Change.pdf](https://www.wetlands.org/wp-content/uploads/2015/11/Wetlands-and-Global-Climate-Change.pdf) (Accessed June 19, 2018).
- 39 van der Esch, S., and Coauthors, 2017: *Exploring future changes in land use and land condition and*
40 *the impacts on food, water, climate change and biodiversity: Scenarios for the Global Land*
41 *Outlook*. The Hague, The Netherlands, [https://www.pbl.nl/sites/default/files/cms/publicaties/pbl-](https://www.pbl.nl/sites/default/files/cms/publicaties/pbl-2017-exploring-future-changes-in-land-use-and-land-condition-2076b.pdf)
42 [2017-exploring-future-changes-in-land-use-and-land-condition-2076b.pdf](https://www.pbl.nl/sites/default/files/cms/publicaties/pbl-2017-exploring-future-changes-in-land-use-and-land-condition-2076b.pdf).
- 43 Etemadi, H., S. Z. Samadi, M. Sharifikia, and J. M. Smoak, 2016: Assessment of climate change
44 downscaling and non-stationarity on the spatial pattern of a mangrove ecosystem in an arid
45 coastal region of southern Iran. *Theor. Appl. Climatol.*, **126**, 35–49, doi:10.1007/s00704-015-
46 1552-5.
- 47 eThekwini Municipal Council, 2014: *The Durban Climate Change Strategy. Environmental Planning*
48 *and Climate Protection Department (EPCPD) and the Energy Office (EO) of eThekwini*

- 1 *Municipality, Durban, South Africa.*
- 2 European Union, 2015: Towards an EU Research and Innovation policy agenda for Nature-Based
3 Solutions & Re-Naturing Cities. *Nature-Based Solut. Re-Naturing Cities* , doi:10.2777/765301.
- 4 Evans, C. D., F. Renou-Wilson, and M. Strack, 2016: The role of waterborne carbon in the
5 greenhouse gas balance of drained and re-wetted peatlands. *Aquat. Sci.*, doi:10.1007/s00027-
6 015-0447-y.
- 7 Evans, R., and J. Boardman, 2016a: A reply to panagos et al., 2016 (Environmental science &
8 policy 59 (2016) 53–57. *Environ. Sci. Policy*, **60**, 63–68, doi:10.1016/J.ENVSCI.2016.03.004.
9 <https://www.sciencedirect.com/science/article/pii/S1462901116300557> (Accessed February 20,
10 2019).
- 11 —, and —, 2016b: The new assessment of soil loss by water erosion in Europe. Panagos P. et al.,
12 2015 Environmental Science & Policy 54, 438–447—A response. *Environ. Sci. Policy*, **58**,
13 11–15, doi:10.1016/j.envsci.2015.12.013.
14 <http://www.sciencedirect.com/science/article/pii/S1462901115301283> (Accessed October 19,
15 2017).
- 16 Evans, S. E., and M. D. Wallenstein, 2014: Climate change alters ecological strategies of soil bacteria.
17 *Ecol. Lett.*, **17**, 155–164, doi:10.1111/ele.12206.
18 <http://www.ncbi.nlm.nih.gov/pubmed/24261594> (Accessed May 2, 2018).
- 19 Fairhead, J., and I. Scoones, 2005: Local knowledge and the social shaping of soil investments:
20 critical perspectives on the assessment of soil degradation in Africa. *Land use policy*, **22**, 33–41,
21 doi:10.1016/J.LANDUSEPOL.2003.08.004.
22 <https://www.sciencedirect.com/science/article/pii/S0264837704000183> (Accessed April 28,
23 2018).
- 24 Fang, Y., B. P. B. Singh, and B. P. B. Singh, 2015: Effect of temperature on biochar priming effects
25 and its stability in soils. *Soil Biol. Biochem.*, **80**, 136–145, doi:10.1016/j.soilbio.2014.10.006.
26 <https://www.sciencedirect.com/science/article/pii/S0038071714003496> (Accessed December 2,
27 2017).
- 28 —, B. P. Singh, P. Matta, A. L. Cowie, and L. Van Zwieten, 2017: Temperature sensitivity and
29 priming of organic matter with different stabilities in a Vertisol with aged biochar. *Soil Biol.*
30 *Biochem.*, **115**, 346–356, doi:10.1016/j.soilbio.2017.09.004.
31 <https://www.sciencedirect.com/science/article/pii/S0038071717304108> (Accessed December 2,
32 2017).
- 33 FAO, 2007: *Land evaluation: towards a revised framework. Land and Water Discussion Paper No. 6.*
34 Rome, Italy,.
- 35 —, 2015: *FRA 2015 Terms and Definitions.* Rome, 1-81 pp.
36 <http://www.fao.org/3/ap862e/ap862e00.pdf>.
- 37 —, 2016: *Global forest resources assessment 2015: how are the world's forests changing?* K.
38 MacDicken, Ö. Jonsson, L. Pina, and S. Maulo, Eds. Food and Agricultural Organization of the
39 UN, Rome, 44 pp. <http://agris.fao.org/agris-search/search.do?recordID=XF2017001127>
40 (Accessed February 25, 2019).
- 41 —, 2018: *From reference levels to results reporting: REDD+ under the UNFCCC, update.* Rome,
42 italy, <http://www.fao.org/3/CA0176EN/ca0176en.pdf>.
- 43 Fargione, J. E., and Coauthors, 2018: Natural climate solutions for the United States. *Sci. Adv.*, **4**,
44 eaat1869, doi:10.1126/sciadv.aat1869.
45 <http://advances.sciencemag.org/content/4/11/eaat1869.abstract>.
- 46 Fasullo, J. T., and R. S. Nerem, 2018: Altimeter-era emergence of the patterns of forced sea-level rise
47 in climate models and implications for the future. *Proc. Natl. Acad. Sci. U. S. A.*, **115**, 12944–
48 12949, doi:10.1073/pnas.1813233115. <http://www.ncbi.nlm.nih.gov/pubmed/30509991>

- 1 (Accessed March 15, 2019).
- 2 Federici, S., F. N. Tubiello, M. Salvatore, H. Jacobs, and J. Schmidhuber, 2015: Forest Ecology and
3 Management New estimates of CO₂ forest emissions and removals : 1990 – 2015. *For. Ecol.*
4 *Manage.*, **352**, 89–98, doi:10.1016/j.foreco.2015.04.022.
5 <http://dx.doi.org/10.1016/j.foreco.2015.04.022>.
- 6 Felton, A., M. Lindbladh, J. Brunet, and Ö. Fritz, 2010: Replacing coniferous monocultures with
7 mixed-species production stands: An assessment of the potential benefits for forest biodiversity
8 in northern Europe. *For. Ecol. Manage.*, **260**, 939–947, doi:10.1016/j.foreco.2010.06.011.
9 <http://dx.doi.org/10.1016/j.foreco.2010.06.011>.
- 10 Fenner, N., R. Williams, H. Toberman, S. Hughes, B. Reynolds, and C. Freeman, 2011:
11 Decomposition ‘hotspots’ in a rewetted peatland: implications for water quality and carbon
12 cycling. *Hydrobiologia*, **674**, 51–66, doi:10.1007/s10750-011-0733-1.
13 <http://link.springer.com/10.1007/s10750-011-0733-1> (Accessed February 14, 2018).
- 14 Fensholt, R., and Coauthors, 2015: Assessing Drivers of Vegetation Changes in Drylands from Time
15 Series of Earth Observation Data. Springer, Cham, 183–202
16 http://link.springer.com/10.1007/978-3-319-15967-6_9 (Accessed October 25, 2018).
- 17 Fernandes, K., L. Verchot, W. Baethgen, V. Gutierrez-Velez, M. Pinedo-Vasquez, and C. Martius,
18 2017: Heightened fire probability in Indonesia in non-drought conditions: The effect of
19 increasing temperatures. *Environ. Res. Lett.*, **12**, doi:10.1088/1748-9326/aa6884.
- 20 Ferner, J., S. Schmidlein, R. T. Guuroh, J. Lopatin, and A. Linstädter, 2018: Disentangling effects of
21 climate and land-use change on West African drylands’ forage supply. *Glob. Environ. Chang.*,
22 **53**, 24–38, doi:10.1016/J.GLOENVCHA.2018.08.007.
23 <https://www.sciencedirect.com/science/article/pii/S0959378018303340> (Accessed October 10,
24 2018).
- 25 Feyisa, G. L., K. Dons, and H. Meilby, 2014: Efficiency of parks in mitigating urban heat island
26 effect: An example from Addis Ababa. *Landsc. Urban Plan.*, **123**, 87–95,
27 doi:10.1016/j.landurbplan.2013.12.008.
28 <https://linkinghub.elsevier.com/retrieve/pii/S0169204613002399> (Accessed May 17, 2018).
- 29 Field, J. P., and Coauthors, 2010: The ecology of dust. *Front. Ecol. Environ.*, **8**, 423–430,
30 doi:10.1890/090050. <http://doi.wiley.com/10.1890/090050> (Accessed October 2, 2018).
- 31 Filho, W. L., and J. Nalau, 2018: *Limits to Climate Change Adaptation*. W.L. Filho and J. Nalau, Eds.
32 Springer International Publishing, Berlin, Heidelberg, 1-410 pp. <https://doi.org/10.1007/978-3-319-64599-5>.
- 34 Finlayson, C. M., and Coauthors, 2017: Policy considerations for managing wetlands under a
35 changing climate. *Mar. Freshw. Res.*, **68**, 1803, doi:10.1071/MF16244.
36 <http://www.publish.csiro.au/?paper=MF16244> (Accessed April 16, 2019).
- 37 Fischer, E. M., and R. Knutti, 2015: Anthropogenic contribution to global occurrence of heavy-
38 precipitation and high-temperature extremes. *Nat. Clim. Chang.*, **5**, 560–564,
39 doi:10.1038/nclimate2617. <http://www.nature.com/articles/nclimate2617> (Accessed October 19,
40 2018).
- 41 Fischer, F., J. Hauck, R. Brandhuber, E. Weigl, H. Maier, and K. Auerswald, 2016: Spatio-temporal
42 variability of erosivity estimated from highly resolved and adjusted radar rain data
43 (RADOLAN). *Agric. For. Meteorol.*, **223**, 72–80, doi:10.1016/J.AGRFORMET.2016.03.024.
44 <https://www.sciencedirect.com/science/article/pii/S0168192316302143> (Accessed March 8,
45 2019).
- 46 Fischer, G., M. Shah, H. van Velthuisen, and F. Nachtergaele, 2009: Agro-ecological zones
47 assessment. *Land Use, Land Cover and Soil Sciences - Volume III: Land Use Planning*, Eolss
48 Publishers Co., Oxford, UK, 61–81
49 <https://books.google.se/books?hl=sv&lr=&id=zIXTCwAAQBAJ&oi=fnd&pg=PA61&dq=fisch>

- 1 er+shah+velthuizen+nachtergaele&ots=STeVQMCmdi&sig=PuZFjsGvw2fyzxKno3HeMriYVT
2 k&redir_esc=y#v=onepage&q=fischer shah velthuizen nachtergaele&f=false (Accessed May 15,
3 2018).
- 4 Flores-Rentería, D., A. Rincón, F. Valladares, and J. Curiel Yuste, 2016: Agricultural matrix affects
5 differently the alpha and beta structural and functional diversity of soil microbial communities in
6 a fragmented Mediterranean holm oak forest. *Soil Biol. Biochem.*, **92**,
7 doi:10.1016/j.soilbio.2015.09.015.
- 8 Foley, J., G. Asner, M. Costa, ... M. C.-F. in E., and undefined 2007, Amazonia revealed: forest
9 degradation and loss of ecosystem goods and services in the Amazon Basin. *Wiley Online Libr.*,
10 [http://onlinelibrary.wiley.com/doi/10.1890/1540-](http://onlinelibrary.wiley.com/doi/10.1890/1540-9295(2007)5%5B25:ARFDAL%5D2.0.CO;2/full)
11 [9295\(2007\)5%5B25:ARFDAL%5D2.0.CO;2/full](http://onlinelibrary.wiley.com/doi/10.1890/1540-9295(2007)5%5B25:ARFDAL%5D2.0.CO;2/full) (Accessed December 28, 2017).
- 12 Foley, J. A., and Coauthors, 2005: Global consequences of land use. *Science*, **309**, 570–574,
13 doi:10.1126/science.1111772. <http://www.ncbi.nlm.nih.gov/pubmed/16040698> (Accessed May
14 11, 2018).
- 15 Foley, J. A., and Coauthors, 2011: Solutions for a cultivated planet. *Nature*, **478**, 337–342,
16 doi:10.1038/nature10452. <http://www.nature.com/articles/nature10452> (Accessed May 24,
17 2018).
- 18 Folke, C., T. Hahn, P. Olsson, and J. Norberg, 2005: ADAPTIVE GOVERNANCE OF SOCIAL-
19 ECOLOGICAL SYSTEMS. *Annu. Rev. Environ. Resour.*, **30**, 441–473,
20 doi:10.1146/annurev.energy.30.050504.144511.
21 <http://www.annualreviews.org/doi/10.1146/annurev.energy.30.050504.144511> (Accessed
22 November 2, 2018).
- 23 ———, S. Carpenter, B. H. Walker, M. Scheffer, T. Chapin, and J. Rockström, 2010: Resilience
24 Thinking: Integrating Resilience, Adaptability and Transformability. *Ecol. Soc.*, **15**.
25 <http://hdl.handle.net/10535/7422>.
- 26 Forest Europe, 2016: Sustainable Forest Management Implementation. *Sustain. For. Manag.*
27 *Implement.*,.
- 28 Forsyth, T., 1996: Science, myth and knowledge: Testing himalayan environmental degradation in
29 Thailand. *Geoforum*, **27**, 375–392, doi:10.1016/S0016-7185(96)00020-6.
30 <https://www.sciencedirect.com/science/article/pii/S0016718596000206> (Accessed April 16,
31 2018).
- 32 Frank, D. C., and Coauthors, 2015: Water-use efficiency and transpiration across European forests
33 during the Anthropocene. *Nat. Clim. Chang.*, **5**, 579–583, doi:10.1038/nclimate2614.
34 <http://www.nature.com/articles/nclimate2614> (Accessed March 14, 2019).
- 35 Frei, M., and K. Becker, 2005: Integrated rice-fish culture: Coupled production saves resources. *Nat.*
36 *Resour. Forum*, **29**, 135–143, doi:10.1111/j.1477-8947.2005.00122.x.
37 <http://doi.wiley.com/10.1111/j.1477-8947.2005.00122.x> (Accessed October 30, 2018).
- 38 French, P. W., 2001: *Coastal Defences: Processes, Problems and Solutions*. Routledge, London.,
- 39 Frenkel, O., A. K. Jaiswal, Y. Elad, B. Lew, C. Kammann, and E. R. Graber, 2017: The effect of
40 biochar on plant diseases: what should we learn while designing biochar substrates? *J. Environ.*
41 *Eng. Landsc. Manag.*, doi:10.3846/16486897.2017.1307202.
- 42 Friend, A. D., and Coauthors, 2014: Carbon residence time dominates uncertainty in terrestrial
43 vegetation responses to future climate and atmospheric CO₂. *Proc. Natl. Acad. Sci. U. S. A.*,
44 **111**, 3280–3285, doi:10.1073/pnas.1222477110.
45 <http://www.ncbi.nlm.nih.gov/pubmed/24344265> (Accessed October 24, 2018).
- 46 Fritsche, U. R., and Coauthors, 2017: Energy and Land Use.
- 47 Fritz, H. M., C. D. Blount, S. Thwin, M. K. Thu, and N. Chan, 2009: Cyclone Nargis storm surge in
48 Myanmar. *Nat. Geosci.*, **2**, 448–449, doi:10.1038/ngeo558.

- 1 <http://www.nature.com/articles/ngeo558> (Accessed April 10, 2019).
- 2 ———, C. Blount, S. Thwin, M. K. Thu, and N. Chan, 2010: Cyclone Nargis Storm Surge Flooding in
3 Myanmar's Ayeyarwady River Delta. *Indian Ocean Tropical Cyclones and Climate Change*,
4 Springer Netherlands, Dordrecht, 295–303 http://www.springerlink.com/index/10.1007/978-90-481-3109-9_34 (Accessed April 10, 2019).
- 6 Frothing, S., J. Talbot, M. C. Jones, C. C. Treat, J. B. Kauffman, E.-S. Tuittila, and N. Roulet, 2011:
7 Peatlands in the Earth's 21st century climate system. *Environ. Rev.*, **19**, 371–396,
8 doi:10.1139/a11-014.
- 9 Froude, M. J., and D. N. Petley, 2018: Global fatal landslide occurrence from 2004 to 2016. *Nat.*
10 *Hazards Earth Syst. Sci.*, **18**, 2161–2181, doi:10.5194/nhess-18-2161-2018. [https://www.nat-](https://www.nat-hazards-earth-syst-sci.net/18/2161/2018/)
11 [hazards-earth-syst-sci.net/18/2161/2018/](https://www.nat-hazards-earth-syst-sci.net/18/2161/2018/) (Accessed April 10, 2019).
- 12 Fryd, O., S. Pauleit, and O. Bühler, 2011: The role of urban green space and trees in relation to
13 climate change. *CAB Rev. Perspect. Agric. Vet. Sci. Nutr. Nat. Resour.*, **6**, 1–18,
14 doi:10.1079/PAVSNR20116053. <https://www.cabdirect.org/cabdirect/abstract/20113401203>
15 (Accessed May 17, 2018).
- 16 Fuangwasdi, A., S. Worakijthamrong, and S. D. Shah, 2019: Addressing Subsidence in Bangkok,
17 Thailand and Houston, Texas: Scientific Comparisons and Data-Driven Groundwater Policies
18 for Coastal Land-Surface Subsidence. *IAEG/AEG Annual Meeting Proceedings, San Francisco,*
19 *California, 2018 - Volume 5*, Springer International Publishing, Cham, 51–60
20 http://link.springer.com/10.1007/978-3-319-93136-4_7 (Accessed March 17, 2019).
- 21 Fujiki, S., and Coauthors, 2016: Large-Scale Mapping of Tree-Community Composition as a
22 Surrogate of Forest Degradation in Bornean Tropical Rain Forests. *Land*, **5**, 45,
23 doi:10.3390/land5040045. <http://www.mdpi.com/2073-445X/5/4/45> (Accessed March 18,
24 2019).
- 25 Fuller, D. Q., and Coauthors, 2011: The contribution of rice agriculture and livestock pastoralism to
26 prehistoric methane levels. *The Holocene*, **21**, 743–759, doi:10.1177/0959683611398052.
27 <http://journals.sagepub.com/doi/10.1177/0959683611398052> (Accessed October 2, 2018).
- 28 Fuso Nerini, F., C. Ray, and Y. Boulkaid, 2017: The cost of cooking a meal. The case of Nyeri
29 County, Kenya. *Environ. Res. Lett.*, **12**, 65007, doi:10.1088/1748-9326/aa6fd0.
30 <http://dx.doi.org/10.1088/1748-9326/aa6fd0>.
- 31 Fuss, S., and Coauthors, 2018: Negative emissions — Part 2: Costs , potentials and side effects
32 OPEN ACCESS Negative emissions — Part 2 : Costs , potentials and side effects.
- 33 Gabrielsson, S., S. Brogaard, and A. Jerneck, 2013: Living without buffers—illustrating climate
34 vulnerability in the Lake Victoria basin. *Sustain. Sci.*, **8**, 143–157, doi:10.1007/s11625-012-
35 0191-3. <http://link.springer.com/10.1007/s11625-012-0191-3> (Accessed May 16, 2018).
- 36 Gamache, I., and S. Payette, 2005: Latitudinal Response of Subarctic Tree Lines to Recent Climate
37 Change in Eastern Canada. *J. Biogeogr.*, **32**, 849–862.
- 38 Gao, B., A. R. Taylor, E. B. Searle, P. Kumar, Z. Ma, A. M. Hume, and H. Y. H. Chen, 2018: Carbon
39 Storage Declines in Old Boreal Forests Irrespective of Succession Pathway. *Ecosystems*, **21**, 1–
40 15, doi:10.1007/s10021-017-0210-4.
- 41 Gao, Q., M. Zribi, M. Escorihuela, N. Baghdadi, Q. Gao, M. Zribi, M. J. Escorihuela, and N.
42 Baghdadi, 2017: Synergetic Use of Sentinel-1 and Sentinel-2 Data for Soil Moisture Mapping at
43 100 m Resolution. *Sensors*, **17**, 1966, doi:10.3390/s17091966. <http://www.mdpi.com/1424-8220/17/9/1966> (Accessed October 1, 2018).
- 45 Garbrecht, J. D., and X. C. Zhang, 2015: Soil Erosion from Winter Wheat Cropland under Climate
46 Change in Central Oklahoma. *Appl. Eng. Agric.*, **31**, 439–454, doi:10.13031/aea.31.10998.
47 <http://elibrary.asabe.org/abstract.asp?aid=45622&t=3&dabs=Y&redir=&redirType=> (Accessed
48 May 23, 2018).

- 1 ———, J. L. Steiner, and A. Cox, Craig, 2007: The times they are changing: soil and water conservation
2 in the 21st century. *Hydrol. Process.*, **21**, 2677–2679.
- 3 García-Ruiz, J. M., S. Beguería, E. Nadal-Romero, J. C. González-Hidalgo, N. Lana-Renault, and Y.
4 Sanjuán, 2015: A meta-analysis of soil erosion rates across the world. *Geomorphology*, **239**,
5 160–173, doi:10.1016/j.geomorph.2015.03.008.
6 <http://www.sciencedirect.com/science/article/pii/S0169555X1500149X> (Accessed October 19,
7 2017).
- 8 Gariano, S. L., and F. Guzzetti, 2016: Landslides in a changing climate. *Earth-Science Rev.*, **162**,
9 227–252, doi:10.1016/J.EARSCIREV.2016.08.011.
10 <https://www.sciencedirect.com/science/article/pii/S0012825216302458> (Accessed April 10,
11 2019).
- 12 Gasparatos, A., and Coauthors, 2018: Survey of local impacts of biofuel crop production and adoption
13 of ethanol stoves in southern Africa. *Sci. data*, **5**, 180186, doi:10.1038/sdata.2018.186.
14 <https://www.ncbi.nlm.nih.gov/pubmed/30226483>.
- 15 Gauthier, S., P. Bernier, T. Kuuluvainen, A. Z. Shvidenko, and D. G. Schepaschenko, 2015: Boreal
16 forest health and global change. *Science*, **349**, 819–822, doi:10.1126/science.aaa9092.
17 <http://www.ncbi.nlm.nih.gov/pubmed/26293953> (Accessed October 27, 2018).
- 18 Gaveau, D. L. a, and Coauthors, 2014: Major atmospheric emissions from peat fires in Southeast Asia
19 during non-drought years: evidence from the 2013 Sumatran fires. *Sci. Rep.*, **4**, 1–7,
20 doi:10.1038/srep06112.
- 21 Ge, J., and C. Zou, 2013: Impacts of woody plant encroachment on regional climate in the southern
22 Great Plains of the United States. *J. Geophys. Res. Atmos.*, **118**, 9093–9104,
23 doi:10.1002/jgrd.50634. <http://doi.wiley.com/10.1002/jgrd.50634> (Accessed October 3, 2018).
- 24 Gebara, M., and A. Agrawal, 2017:
25 Beyond Rewards and Punishments in the Brazilian Amazon: Practical Implications of the RED
26 D+ Discourse. *Forests*, **8**, 66, doi:10.3390/f8030066. <http://www.mdpi.com/1999-4907/8/3/66>
27 (Accessed May 27, 2018).
- 28 Gebara, M. F., 2018: Tenure reforms in indigenous lands: decentralized forest management or
29 illegalism? *Curr. Opin. Environ. Sustain.*, **32**, 60–67, doi:10.1016/J.COSUST.2018.04.008.
30 <https://www.sciencedirect.com/science/article/pii/S1877343517301641> (Accessed October 19,
31 2018).
- 32 Geels, F. W., 2002: Technological transitions as evolutionary reconfiguration processes: a multi-level
33 perspective and a case-study. *Res. Policy*, **31**, 1257–1274, doi:10.1016/S0048-7333(02)00062-8.
34 <https://www.sciencedirect.com/science/article/pii/S0048733302000628> (Accessed October 5,
35 2018).
- 36 Van der Geest, K., and M. Schindler, 2016: *Case study report: Loss and damage from a catastrophic*
37 *landslide in Sindhupalchok District, Nepal. Report No.1.* Bonn, Germany, 1-96 pp.
38 https://www.raonline.ch/pages/np/pdf/UNU-EHS_Nepalstudy2016.pdf.
- 39 GEF, 2018: *GEF-7 Replenishment, Programming Directions.* Stockholm, Sweden,
40 [https://www.thegef.org/sites/default/files/publications/GEF-7_Programming_Directions -
41 GEF_R.7_19.pdf](https://www.thegef.org/sites/default/files/publications/GEF-7_Programming_Directions_-_GEF_R.7_19.pdf).
- 42 Genesisio, L., F. Miglietta, E. Lugato, S. Baronti, M. Pieri, and F. P. Vaccari, 2012: Surface albedo
43 following biochar application in durum wheat. *Environ. Res. Lett.*, doi:10.1088/1748-
44 9326/7/1/014025.
- 45 Genesisio, L., F. P. Vaccari, and F. Miglietta, 2016: Black carbon aerosol from biochar threatens its
46 negative emission potential. *Glob. Chang. Biol.*, doi:10.1111/gcb.13254.
- 47 Gerber, N., E. Nkonya, and J. von Braun, 2014: Land Degradation, Poverty and Marginality.
48 *Marginality*, Springer Netherlands, Dordrecht, 181–202 <http://link.springer.com/10.1007/978->

- 1 94-007-7061-4_12 (Accessed March 5, 2019).
- 2 Gerten, D., R. Betts, and P. Döll, 2014: Cross-chapter box on the active role of vegetation in altering
3 water flows under climate change. *Climate Change 2014: Impacts, Adaptation, and*
4 *Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the*
5 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, C.B. Field et al.,
6 Eds., Cambridge University Press, Cambridge, UK and New York, USA, 157–161.
- 7 van Gestel, N., and Coauthors, 2018: Predicting soil carbon loss with warming. *Nature*, **554**, E4–E5,
8 doi:10.1038/nature25745. <http://www.nature.com/doi/10.1038/nature25745> (Accessed
9 February 27, 2019).
- 10 Gharbaoui, D., and J. Blocher, 2016: The Reason Land Matters: Relocation as Adaptation to Climate
11 Change in Fiji Islands. Springer, Cham, 149–173 [http://link.springer.com/10.1007/978-3-319-](http://link.springer.com/10.1007/978-3-319-42922-9_8)
12 [42922-9_8](http://link.springer.com/10.1007/978-3-319-42922-9_8) (Accessed March 30, 2019).
- 13 Ghazoul, J., and R. Chazdon, 2017: Degradation and Recovery in Changing Forest Landscapes: A
14 Multiscale Conceptual Framework. *Annu. Rev. Environ. Resour.*, **42**, 161–188,
15 doi:10.1146/annurev-environ. [http://www.annualreviews.org/doi/10.1146/annurev-environ-](http://www.annualreviews.org/doi/10.1146/annurev-environ-102016-060736)
16 [102016-060736](http://www.annualreviews.org/doi/10.1146/annurev-environ-102016-060736) (Accessed November 2, 2017).
- 17 —, Z. Burivalova, J. Garcia-Ulloa, and L. A. King, 2015: Conceptualizing Forest Degradation.
18 *Trends Ecol. Evol.*, **30**, 622–632, doi:10.1016/j.tree.2015.08.001.
19 <http://www.ncbi.nlm.nih.gov/pubmed/26411619> (Accessed October 1, 2018).
- 20 Ghosh, A., S. Schmidt, T. Fickert, and M. Nüsser, 2015: The Indian Sundarban Mangrove Forests:
21 History, Utilization, Conservation Strategies and Local Perception. *Diversity*, **7**, 149–169,
22 doi:10.3390/d7020149. <http://www.mdpi.com/1424-2818/7/2/149> (Accessed June 19, 2018).
- 23 Gibbs, H. K., and J. M. Salmon, 2015: Mapping the world’s degraded lands. *Appl. Geogr.*, **57**, 12–21,
24 doi:10.1016/J.APGEOG.2014.11.024.
25 <https://www.sciencedirect.com/science/article/pii/S0143622814002793> (Accessed November 2,
26 2017).
- 27 Gideon Neba, S., M. Kanninen, R. Eba’a Atyi, and D. J. Sonwa, 2014: Assessment and prediction of
28 above-ground biomass in selectively logged forest concessions using field measurements and
29 remote sensing data: Case study in South East Cameroon. *For. Ecol. Manage.*, **329**, 177–185,
30 doi:10.1016/J.FORECO.2014.06.018.
31 <https://www.sciencedirect.com/science/article/pii/S037811271400382X> (Accessed October 31,
32 2018).
- 33 Giger, M., H. Liniger, C. Sauter, and G. Schwilch, 2018: Economic Benefits and Costs of Sustainable
34 Land Management Technologies: An Analysis of WOCAT’s Global Data. *L. Degrad. Dev.*, **29**,
35 962–974, doi:10.1002/ldr.2429. <http://doi.wiley.com/10.1002/ldr.2429> (Accessed April 17,
36 2019).
- 37 Giguère-Croteau, C., É. Boucher, Y. Bergeron, M. P. Girardin, I. Drobyshev, L. C. R. Silva, J.-F.
38 Hélie, and M. Garneau, 2019: North America’s oldest boreal trees are more efficient water users
39 due to increased [CO₂], but do not grow faster. *Proc. Natl. Acad. Sci.*, **116**, 2749–2754,
40 doi:10.1073/pnas.1816686116.
- 41 Gill, S. ., J. . Handley, A. . Ennos, and S. Pauleit, 2007: Adapting Cities for Climate Change: The
42 Role of the Green Infrastructure. *Built Environ.*, **33**, 115–133, doi:10.2148/benv.33.1.115.
43 [http://openurl.ingenta.com/content/xref?genre=article&issn=0263-](http://openurl.ingenta.com/content/xref?genre=article&issn=0263-7960&volume=33&issue=1&spage=115)
44 [7960&volume=33&issue=1&spage=115](http://openurl.ingenta.com/content/xref?genre=article&issn=0263-7960&volume=33&issue=1&spage=115) (Accessed May 17, 2018).
- 45 Gingrich, S., and Coauthors, 2015: Exploring long-term trends in land use change and aboveground
46 human appropriation of net primary production in nine European countries. *Land use policy*, **47**,
47 426–438, doi:10.1016/J.LANDUSEPOL.2015.04.027.
48 <https://www.sciencedirect.com/science/article/pii/S0264837715001374> (Accessed October 31,
49 2018).

- 1 Giorgi, F., and P. Lionello, 2008: Climate change projections for the Mediterranean region. *Glob.*
2 *Planet. Change*, **63**, 90–104, doi:10.1016/J.GLOPLACHA.2007.09.005.
3 <https://www.sciencedirect.com/science/article/pii/S0921818107001750> (Accessed October 24,
4 2018).
- 5 Girardin, M. P., and Coauthors, 2016: No growth stimulation of Canada's boreal forest under half-
6 century of combined warming and CO₂ fertilization. *Proc. Natl. Acad. Sci. U. S. A.*, **113**,
7 E8406–E8414, doi:10.1073/pnas.1610156113. <http://www.ncbi.nlm.nih.gov/pubmed/27956624>
8 (Accessed October 27, 2018).
- 9 Glover, J. D., and Coauthors, 2010: Harvested perennial grasslands provide ecological benchmarks
10 for agricultural sustainability. *Agric. Ecosyst. Environ.*, **137**, 3–12,
11 doi:10.1016/j.agee.2009.11.001.
- 12 Gockowski, J., and D. Sonwa, 2011: Cocoa intensification scenarios and their predicted impact on CO
13 2 emissions, biodiversity conservation, and rural livelihoods in the Guinea rain forest of West
14 Africa. *Environ. Manage.*, **48**, 307–321, doi:10.1007/s00267-010-9602-3.
- 15 Gonzalez-Hidalgo, J. C., M. de Luis, and R. J. Batalla, 2009: Effects of the largest daily events on
16 total soil erosion by rainwater. An analysis of the USLE database. *Earth Surf. Process.*
17 *Landforms*, **34**, 2070–2077, doi:10.1002/esp.1892. <http://doi.wiley.com/10.1002/esp.1892>
18 (Accessed March 9, 2018).
- 19 Gonzalez-Hidalgo, J. C., R. J. Batalla, A. Cerda, and M. de Luis, 2012: A regional analysis of the
20 effects of largest events on soil erosion. *CATENA*, **95**, 85–90,
21 doi:10.1016/J.CATENA.2012.03.006.
22 <https://www.sciencedirect.com/science/article/pii/S0341816212000616?via%3Dihub> (Accessed
23 March 9, 2018).
- 24 Gonzalez, P., R. P. Neilson, J. M. Lenihan, and R. J. Drapek, 2010: Global patterns in the
25 vulnerability of ecosystems to vegetation shifts due to climate change. *Glob. Ecol. Biogeogr.*,
26 **19**, 755–768, doi:10.1111/j.1466-8238.2010.00558.x. <http://doi.wiley.com/10.1111/j.1466-8238.2010.00558.x> (Accessed March 18, 2019).
- 28 Gopal, B., and M. Chauhan, 2006: Biodiversity and its conservation in the Sundarban Mangrove
29 Ecosystem. *Aquat. Sci.*, **68**, 338–354, doi:10.1007/s00027-006-0868-8.
30 <http://link.springer.com/10.1007/s00027-006-0868-8> (Accessed June 19, 2018).
- 31 Government of Tuvalu, 2006: *National Action Plan to Combat Land Degradation and Drought*.
32 Funafuti,.
- 33 Grainger, A., 2009: The role of science in implementing international environmental agreements: The
34 case of desertification. *L. Degrad. Dev.*, **20**, 410–430, doi:10.1002/ldr.898.
35 <http://doi.wiley.com/10.1002/ldr.898> (Accessed May 12, 2018).
- 36 Grandy, A. S., and G. P. Robertson, 2006: Aggregation and Organic Matter Protection Following
37 Tillage of a Previously Uncultivated Soil. *Soil Sci. Soc. Am. J.*, **70**, 1398,
38 doi:10.2136/sssaj2005.0313. <https://www.soils.org/publications/sssaj/abstracts/70/4/1398>
39 (Accessed May 21, 2018).
- 40 —, and J. C. Neff, 2008: Molecular C dynamics downstream: The biochemical decomposition
41 sequence and its impact on soil organic matter structure and function. *Sci. Total Environ.*, **404**,
42 297–307, doi:10.1016/j.scitotenv.2007.11.013.
- 43 Grassini, P., K. M. Eskridge, and K. G. Cassman, 2013: Distinguishing between yield advances and
44 yield plateaus in historical crop production trends. *Nat. Commun.*, **4**, 2918,
45 doi:10.1038/ncomms3918. <http://www.nature.com/articles/ncomms3918> (Accessed May 23,
46 2018).
- 47 Gray, C., and R. Bilborrow, 2013: Environmental Influences on Human Migration in Rural Ecuador.
48 *Demography*, **50**, 1217–1241, doi:10.1007/s13524-012-0192-y.
49 <http://link.springer.com/10.1007/s13524-012-0192-y> (Accessed May 15, 2018).

- 1 Gray, C. L., 2011: Soil quality and human migration in Kenya and Uganda. *Glob. Environ. Chang.*,
2 **21**, 421–430, doi:10.1016/J.GLOENVCHA.2011.02.004.
3 <https://www.sciencedirect.com/science/article/pii/S0959378011000264> (Accessed March 11,
4 2019).
- 5 Green, A. J., and Coauthors, 2017: Creating a safe operating space for wetlands in a changing climate.
6 *Front. Ecol. Environ.*, **15**, 99–107, doi:10.1002/fee.1459. <http://doi.wiley.com/10.1002/fee.1459>
7 (Accessed March 18, 2019).
- 8 Green Surge, 2016: *Advancing approaches and strategies for UGI planning and implementation.*
9 *Green Surge Deliverable 5.2.* Available online: [http://greensurge.eu/working-](http://greensurge.eu/working-packages/wp5/files/D5_2_Hansen_et_al_2016_Advanced_UGI_Planning_and_Implementation_v3.pdf)
10 [packages/wp5/files/D5_2_Hansen_et_al_2016_Advanced_UGI_Planning_and_Implementation_](http://greensurge.eu/working-packages/wp5/files/D5_2_Hansen_et_al_2016_Advanced_UGI_Planning_and_Implementation_v3.pdf)
11 [v3.pdf](http://greensurge.eu/working-packages/wp5/files/D5_2_Hansen_et_al_2016_Advanced_UGI_Planning_and_Implementation_v3.pdf).
- 12 Grimm, V., and J. M. Calabrese, 2011: What Is Resilience? A Short Introduction. Springer, Berlin,
13 Heidelberg, 3–13 http://link.springer.com/10.1007/978-3-642-20423-4_1 (Accessed March 30,
14 2019).
- 15 Griscom, B. W., and Coauthors, 2017: Natural climate solutions. *Proc. Natl. Acad. Sci.*, **114**, 11645–
16 11650, doi:10.1073/pnas.1710465114.
17 <http://www.pnas.org/content/early/2017/10/11/1710465114.full.pdf> (Accessed October 21,
18 2017).
- 19 van Groenigen, J. W., C. van Kessel, B. A. Hungate, O. Oenema, D. S. Powlson, and K. J. van
20 Groenigen, 2017: Sequestering Soil Organic Carbon: A Nitrogen Dilemma. *Environ. Sci.*
21 *Technol.*, **51**, 4738–4739, doi:10.1021/acs.est.7b01427.
22 <http://pubs.acs.org/doi/10.1021/acs.est.7b01427> (Accessed March 29, 2019).
- 23 Groisman, P. Y., and Coauthors, 2005: Trends in Intense Precipitation in the Climate Record. *J.*
24 *Clim.*, **18**, 1326–1350, doi:10.1175/JCLI3339.1.
25 <http://journals.ametsoc.org/doi/abs/10.1175/JCLI3339.1> (Accessed March 7, 2019).
- 26 Grubler, A., and Coauthors, 2018: A low energy demand scenario for meeting the 1.5 °C target and
27 sustainable development goals without negative emission technologies. *Nat. Energy*, **3**, 515–
28 527, doi:10.1038/s41560-018-0172-6. <http://www.nature.com/articles/s41560-018-0172-6>
29 (Accessed April 10, 2019).
- 30 Guerreiro, S. B., H. J. Fowler, R. Barbero, S. Westra, G. Lenderink, S. Blenkinsop, E. Lewis, and X.-
31 F. Li, 2018: Detection of continental-scale intensification of hourly rainfall extremes. *Nat. Clim.*
32 *Chang.*, **8**, 803–807, doi:10.1038/s41558-018-0245-3. [http://www.nature.com/articles/s41558-](http://www.nature.com/articles/s41558-018-0245-3)
33 [018-0245-3](http://www.nature.com/articles/s41558-018-0245-3) (Accessed October 22, 2018).
- 34 Gumbricht, T., R. M. Roman-Cuesta, L. Verchot, M. Herold, F. Wittmann, E. Householder, N.
35 Herold, and D. Murdiyoso, 2017: An expert system model for mapping tropical wetlands and
36 peatlands reveals South America as the largest contributor. *Glob. Chang. Biol.*, **23**, 3581–3599,
37 doi:10.1111/gcb.13689.
- 38 Guo, J. H., and Coauthors, 2010: Significant acidification in major Chinese croplands. *Science*, **327**,
39 1008–1010, doi:10.1126/science.1182570. <http://www.ncbi.nlm.nih.gov/pubmed/20150447>
40 (Accessed October 1, 2018).
- 41 Guo, L. B., and R. M. Gifford, 2002: Soil carbon stocks and land use change: a meta analysis. *Glob.*
42 *Chang. Biol.*, **8**, 345–360, doi:10.1046/j.1354-1013.2002.00486.x.
- 43 Gupta, S., and S. Kumar, 2017: Simulating climate change impact on soil erosion using RUSLE
44 model – A case study in a watershed of mid-Himalayan landscape. *J. Earth Syst. Sci.*, **126**, 43,
45 doi:10.1007/s12040-017-0823-1. <http://link.springer.com/10.1007/s12040-017-0823-1>
46 (Accessed March 13, 2019).
- 47 Haberl, H., K.-H. Erb, F. Krausmann, S. Running, T. D. Searchinger, and W. Kolby Smith, 2013:
48 Bioenergy: how much can we expect for 2050? *Environ. Res. Lett.*, **8**, 031004,
49 doi:10.1088/1748-9326/8/3/031004. [Subject to Copy-editing](http://stacks.iop.org/1748-</p></div><div data-bbox=)

- 1 9326/8/i=3/a=031004?key=crossref.6aa16b28ce5251381d8db735788049b2 (Accessed April 11,
2 2019).
- 3 Hadden, D., and A. Grelle, 2016: Changing temperature response of respiration turns boreal forest
4 from carbon sink into carbon source. *Agric. For. Meteorol.*, **223**, 30–38,
5 doi:10.1016/j.agrformet.2016.03.020.
- 6 Haider, G., D. Steffens, G. Moser, C. Müller, and C. I. Kammann, 2017: Biochar reduced nitrate
7 leaching and improved soil moisture content without yield improvements in a four-year field
8 study. *Agric. Ecosyst. Environ.*, doi:10.1016/j.agee.2016.12.019.
- 9 Håkansson, N. T., and M. Widgren, 2007: Labour and landscapes: the political economy of landesque
10 capital in nineteenth century tanganyika. *Geogr. Ann. Ser. B, Hum. Geogr.*, **89**, 233–248,
11 doi:10.1111/j.1468-0467.2007.00251.x. [https://www.tandfonline.com/doi/full/10.1111/j.1468-](https://www.tandfonline.com/doi/full/10.1111/j.1468-0467.2007.00251.x)
12 [0467.2007.00251.x](https://www.tandfonline.com/doi/full/10.1111/j.1468-0467.2007.00251.x) (Accessed March 27, 2019).
- 13 Hallegatte, S., M. Fay, and E. B. Barbier, 2018: Poverty and climate change: introduction. *Environ.*
14 *Dev. Econ.*, **23**, 217–233, doi:10.1017/S1355770X18000141.
15 https://www.cambridge.org/core/product/identifier/S1355770X18000141/type/journal_article
16 (Accessed March 8, 2019).
- 17 Halvorson, W. L., A. E. Castellanos, and J. Murrieta-Saldivar, 2003: Sustainable Land Use Requires
18 Attention to Ecological Signals. *Environ. Manage.*, **32**, 551–558, doi:10.1007/s00267-003-2889-
19 6.
- 20 Hamza, M. A., and W. K. Anderson, 2005: Soil compaction in cropping systems. *Soil Tillage Res.*,
21 **82**, 121–145, doi:10.1016/j.still.2004.08.009.
22 <http://linkinghub.elsevier.com/retrieve/pii/S0167198704001849> (Accessed February 14, 2018).
- 23 Handmer, J., and J. Nalau, 2019: Understanding Loss and Damage in Pacific Small Island Developing
24 States. Springer, Cham, 365–381 http://link.springer.com/10.1007/978-3-319-72026-5_15
25 (Accessed April 11, 2019).
- 26 Hansen, M. C., and Coauthors, 2013: High-Resolution Global Maps of 21st-Century Forest Cover
27 Change. *Science (80-.)*, **342**, 850–853.
- 28 Haque, F., R. M. Santos, A. Dutta, M. Thimmanagari, and Y. W. Chiang, 2019: Co-Benefits of
29 Wollastonite Weathering in Agriculture: CO₂ Sequestration and Promoted Plant Growth. *ACS*
30 *Omega*, **4**, 1425–1433, doi:10.1021/acsomega.8b02477.
31 <http://pubs.acs.org/doi/10.1021/acsomega.8b02477> (Accessed April 12, 2019).
- 32 Haregeweyn, N., and Coauthors, 2015: Soil erosion and conservation in Ethiopia. *Prog. Phys. Geogr.*
33 *Earth Environ.*, **39**, 750–774, doi:10.1177/0309133315598725.
34 <http://journals.sagepub.com/doi/10.1177/0309133315598725> (Accessed April 3, 2019).
- 35 Harley, M. D., and Coauthors, 2017a: Extreme coastal erosion enhanced by anomalous extratropical
36 storm wave direction. *Sci. Rep.*, **7**, 6033, doi:10.1038/s41598-017-05792-1.
37 <http://www.nature.com/articles/s41598-017-05792-1> (Accessed October 1, 2018).
- 38 —, and Coauthors, 2017b: Extreme coastal erosion enhanced by anomalous extratropical storm
39 wave direction. *Sci. Rep.*, **7**, 6033, doi:10.1038/s41598-017-05792-1.
40 <http://www.nature.com/articles/s41598-017-05792-1> (Accessed March 16, 2019).
- 41 Harmon, M. E., W. K. Ferrell, and J. F. Franklin, 1990: Effects on Carbon Storage of Conversion of
42 Old-Growth Forests to Young Forests. *Science (80-.)*, **247**, 699–702.
- 43 Harper, A. B., and Coauthors, 2018: Land-use emissions play a critical role in land-based mitigation
44 for Paris climate targets. *Nat. Commun.*, **9**, doi:10.1038/s41467-018-05340-z.
- 45 Harsch, M. A., P. E. Hulme, M. S. McGlone, and R. P. Duncan, 2009: Are treelines advancing? A
46 global meta-analysis of treeline response to climate warming. *Ecol. Lett.*, **12**, 1040–1049,
47 doi:10.1111/j.1461-0248.2009.01355.x.

- 1 Hauer, M. E., J. M. Evans, and D. R. Mishra, 2016: Millions projected to be at risk from sea-level rise
2 in the continental United States. *Nat. Clim. Chang.*, **6**, 691–695, doi:10.1038/nclimate2961.
3 <http://www.nature.com/articles/nclimate2961> (Accessed March 14, 2019).
- 4 Hayes, R., and Coauthors, 2018: The Performance of Early-Generation Perennial Winter Cereals at 21
5 Sites across Four Continents. *Sustainability*, **10**, 1124, doi:10.3390/su10041124.
6 <http://www.mdpi.com/2071-1050/10/4/1124> (Accessed May 21, 2018).
- 7 He, Y., and Coauthors, 2017: Effects of biochar application on soil greenhouse gas fluxes: a meta-
8 analysis. *GCB Bioenergy*, **9**, 743–755, doi:10.1111/gcbb.12376.
9 <http://doi.wiley.com/10.1111/gcbb.12376> (Accessed April 2, 2019).
- 10 Hecht, and S. B., 1983: Cattle ranching in the eastern Amazon: Environmental and social
11 implications. *The dilemma of Amazonian Development*, E.F. Moran, Ed., Westview Press,
12 Boulder, CO, USA, 155–188 <https://ci.nii.ac.jp/naid/10021062539/> (Accessed May 15, 2018).
- 13 Heck, V., D. Gerten, W. Lucht, and A. Popp, 2018: Biomass-based negative emissions difficult to
14 reconcile with planetary boundaries. *Nat. Clim. Chang.*, **8**, 151–155, doi:10.1038/s41558-017-
15 0064-y. <http://dx.doi.org/10.1038/s41558-017-0064-y>.
- 16 Hellmann, J. J., J. E. Byers, B. G. Bierwagen, and J. S. Dukes, 2008: Five Potential Consequences of
17 Climate Change for Invasive Species. *Conserv. Biol.*, **22**, 534–543, doi:10.1111/j.1523-
18 1739.2008.00951.x. <http://doi.wiley.com/10.1111/j.1523-1739.2008.00951.x> (Accessed March
19 7, 2019).
- 20 Hember, R. A., W. A. Kurz, and N. C. Coops, 2016: Relationships between individual-tree mortality
21 and water-balance variables indicate positive trends in water stress-induced tree mortality across
22 North America. *Glob. Chang. Biol.*, **23**, 1691–1710, doi:10.1111/gcb.13428.
- 23 —, —, and —, 2017: Increasing net ecosystem biomass production of Canada’s boreal and
24 temperate forests despite decline in dry climates. *Global Biogeochem. Cycles*, **31**, 134–158,
25 doi:10.1002/2016GB005459. <http://doi.wiley.com/10.1002/2016GB005459> (Accessed
26 November 1, 2018).
- 27 Henry, B., B. Murphy, and A. Cowie, 2018: *Sustainable Land Management for Environmental*
28 *Benefits and Food Security A synthesis report for the GEF*. Washington DC, USA, 127 pp.
29 <http://stapgef.org/sites/default/files/publications/SLM20180812.pdf>.
- 30 Henttonen, H. M., P. Nöjd, and H. Mäkinen, 2017: Environment-induced growth changes in the
31 Finnish forests during 1971–2010 – An analysis based on National Forest Inventory. *For. Ecol.*
32 *Manage.*, **386**, 22–36, doi:10.1016/j.foreco.2016.11.044.
33 <http://dx.doi.org/10.1016/j.foreco.2016.11.044>.
- 34 Herbert, E. R., and Coauthors, 2015: A global perspective on wetland salinization: ecological
35 consequences of a growing threat to freshwater wetlands. *Ecosphere*, **6**, art206,
36 doi:10.1890/ES14-00534.1. <http://doi.wiley.com/10.1890/ES14-00534.1> (Accessed December
37 27, 2017).
- 38 Hergoualc’h, K., V. H. Gutiérrez-vélez, M. Menton, and L. V Verchot, 2017a: Forest Ecology and
39 Management Characterizing degradation of palm swamp peatlands from space and on the
40 ground: An exploratory study in the Peruvian Amazon. *For. Ecol. Manage.*, **393**, 63–73,
41 doi:10.1016/j.foreco.2017.03.016.
- 42 Hergoualc’h, K., D. T. Hendry, D. Murdiyarso, and L. V. Verchot, 2017b: Total and heterotrophic
43 soil respiration in a swamp forest and oil palm plantations on peat in Central Kalimantan,
44 Indonesia. *Biogeochemistry*, **135**, 203–220, doi:10.1007/s10533-017-0363-4.
- 45 Hermans, K., and T. Ide, 2019: Advancing Research on Climate Change, Conflict and Migration. *Die*
46 *Erde*, **in press**.
- 47 Hernández-Morcillo, M., T. Plieninger, and C. Bieling, 2013: An empirical review of cultural
48 ecosystem service indicators. *Ecol. Indic.*, **29**, 434–444, doi:10.1016/J.ECOLIND.2013.01.013.

- 1 <https://www.sciencedirect.com/science/article/pii/S1470160X13000320> (Accessed February 28,
2 2019).
- 3 Herrick, J. E., P. Shaver, D. A. Pyke, M. Pellant, D. Toledo, and N. Lepak, 2019: A strategy for
4 defining the reference for land health and degradation assessments. *Ecol. Indic.*, **97**, 225–230,
5 doi:10.1016/J.ECOLIND.2018.06.065.
6 <https://www.sciencedirect.com/science/article/pii/S1470160X18305120> (Accessed February 22,
7 2019).
- 8 Hickey, G. M., M. Pouliot, C. Smith-Hall, S. Wunder, and M. R. Nielsen, 2016: Quantifying the
9 economic contribution of wild food harvests to rural livelihoods: A global-comparative analysis.
10 *Food Policy*, **62**, 122–132, doi:10.1016/J.FOODPOL.2016.06.001.
11 <https://www.sciencedirect.com/science/article/pii/S0306919216300707> (Accessed March 8,
12 2019).
- 13 Hickler, T., L. Eklundh, J. W. Seaquist, B. Smith, J. Ardö, L. Olsson, M. T. Sykes, and M. Sjöström,
14 2005: Precipitation controls Sahel greening trend. *Geophys. Res. Lett.*, **32**, L21415,
15 doi:10.1029/2005GL024370. <http://doi.wiley.com/10.1029/2005GL024370> (Accessed October
16 10, 2018).
- 17 Higginbottom, T., E. Symeonakis, T. P. Higginbottom, and E. Symeonakis, 2014: Assessing Land
18 Degradation and Desertification Using Vegetation Index Data: Current Frameworks and Future
19 Directions. *Remote Sens.*, **6**, 9552–9575, doi:10.3390/rs6109552. <http://www.mdpi.com/2072-4292/6/10/9552> (Accessed September 26, 2018).
- 21 Higgins, S., I. Overeem, A. Tanaka, and J. P. M. Syvitski, 2013: Land subsidence at aquaculture
22 facilities in the Yellow River delta, China. *Geophys. Res. Lett.*, **40**, 3898–3902,
23 doi:10.1002/grl.50758. <http://doi.wiley.com/10.1002/grl.50758> (Accessed October 27, 2018).
- 24 Hilber, I., P. Mayer, V. Gouliarmou, S. E. Hale, G. Cornelissen, H.-P. Schmidt, and T. D. Bucheli,
25 2017: Bioavailability and bioaccessibility of polycyclic aromatic hydrocarbons from (post-
26 pyrolytically treated) biochars. *Chemosphere*, **174**, 700–707,
27 doi:10.1016/J.CHEMOSPHERE.2017.02.014.
28 <https://www.sciencedirect.com/science/article/pii/S0045653517301923> (Accessed April 2,
29 2019).
- 30 Hinsley, A., A. Entwistle, and D. V. Pio, 2015: Does the long-term success of REDD+ also depend on
31 biodiversity? *Oryx*, **49**, 216–221, doi:10.1017/S0030605314000507.
32 http://www.journals.cambridge.org/abstract_S0030605314000507 (Accessed October 18, 2018).
- 33 Hoegh-Guldberg, O., and Coauthors, 2018: Impacts of 1.5°C global warming on natural and human
34 systems. *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of
35 1.5 °C above pre-industrial levels and related global greenhouse gas emission pathways, in the
36 context of strengthening the global response to the threat of climate change*
37 http://report.ipcc.ch/sr15/pdf/sr15_chapter3.pdf (Accessed October 29, 2018).
- 38 Hoffmann, H. K., K. Sander, M. Brüntrup, and S. Sieber, 2017: Applying the Water-Energy-Food
39 Nexus to the Charcoal Value Chain. *Front. Environ. Sci.*, **5**, 84.
40 <https://www.frontiersin.org/article/10.3389/fenvs.2017.00084>.
- 41 Högy, P., and A. Fangmeier, 2008: Effects of elevated atmospheric CO₂ on grain quality of wheat. *J.*
42 *Cereal Sci.*, **48**, 580–591, doi:10.1016/J.JCS.2008.01.006.
43 <https://www.sciencedirect.com/science/article/pii/S0733521008000428> (Accessed May 31,
44 2018).
- 45 Hojas-Gascon, L., H. Eva, D. Ehrlich, M. Pesaresi, A. Frédéric, and J. Garcia, 2016: *Urbanization
46 and forest degradation in east Africa - a case study around Dar es Salaam, Tanzania*. IEEE
47 International Geoscience and Remote Sensing Symposium (IGARSS), 7293-7295 pp.
- 48 Holmberg, M., and Coauthors, 2019: Ecosystem Services Related to Carbon Cycling – Modeling
49 Present and Future Impacts in Boreal Forests. *Front. Plant Sci.*, **10**.

- 1 <https://www.frontiersin.org/article/10.3389/fpls.2019.00343>.
- 2 Holt-Giménez, E., 2002: Measuring farmers' agroecological resistance after Hurricane Mitch in
3 Nicaragua: a case study in participatory, sustainable land management impact monitoring. *Agric.*
4 *Ecosyst. Environ.*, **93**, 87–105, doi:10.1016/S0167-8809(02)00006-3.
5 <https://www.sciencedirect.com/science/article/pii/S0167880902000063> (Accessed November 1,
6 2018).
- 7 Holtum, J. A. M., and K. Winter, 2010: Elevated [CO₂] and forest vegetation: more a water issue
8 than a carbon issue? *Funct. Plant Biol.*, **37**, 694, doi:10.1071/FP10001.
9 <http://www.publish.csiro.au/?paper=FP10001> (Accessed October 24, 2018).
- 10 Homer-Dixon, T. F., J. H. Boutwell, and G. W. Rathjens, 1993: Environmental Change and Violent
11 Conflict. *Sci. Am.*, **268**, 38–45, doi:10.2307/24941373. <http://www.jstor.org/stable/24941373>
12 (Accessed May 15, 2018).
- 13 Hosonuma, N., M. Herold, V. De Sy, R. S. De Fries, M. Brockhaus, L. Verchot, A. Angelsen, and E.
14 Romijn, 2012: An assessment of deforestation and forest degradation drivers in developing
15 countries. *Environ. Res. Lett.*, **7**, 044009, doi:10.1088/1748-9326/7/4/044009.
16 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/7/i=4/a=044009?key=crossref.1a00aa77eac35c904bf7e007011d4763)
17 [9326/7/i=4/a=044009?key=crossref.1a00aa77eac35c904bf7e007011d4763](http://stacks.iop.org/1748-9326/7/i=4/a=044009?key=crossref.1a00aa77eac35c904bf7e007011d4763) (Accessed October
18 31, 2018).
- 19 Hou, B., H. Liao, and J. Huang, 2018: Household cooking fuel choice and economic poverty:
20 Evidence from a nationwide survey in China. *Energy Build.*, **166**, 319–329,
21 doi:10.1016/J.ENBUILD.2018.02.012.
22 <https://www.sciencedirect.com/science/article/pii/S0378778817336241> (Accessed April 10,
23 2019).
- 24 Le Houerou, H. N., 1984: Rain use efficiency: a unifying concept in arid-land ecology. *J. Arid*
25 *Environ.*, **7**, 213–247. <http://agris.fao.org/agris-search/search.do?recordID=US201301473751>
26 (Accessed January 3, 2018).
- 27 Houghton, R. A., and A. A. Nassikas, 2017: Global and regional fluxes of carbon from land use and
28 land cover change 1850-2015. *Global Biogeochem. Cycles*, **31**, 456–472,
29 doi:10.1002/2016GB005546. <http://doi.wiley.com/10.1002/2016GB005546> (Accessed May 26,
30 2018).
- 31 Houghton, R. A., and A. A. Nassikas, 2018: Negative emissions from stopping deforestation and
32 forest degradation, globally. *Glob. Chang. Biol.*, **24**, 350–359, doi:10.1111/gcb.13876.
- 33 Houghton, R. A., J. I. House, J. Pongratz, G. R. Van Der Werf, R. S. Defries, M. C. Hansen, C. Le
34 Quéré, and N. Ramankutty, 2012: Carbon emissions from land use and land-cover change.
35 *Biogeosciences*, **9**, 5125–5142, doi:10.5194/bg-9-5125-2012.
- 36 Houspanossian, J., M. Noretto, and E. G. Jobbágy, 2013: Radiation budget changes with dry forest
37 clearing in temperate Argentina. *Glob. Chang. Biol.*, **19**, 1211–1222, doi:10.1111/gcb.12121.
38 <http://doi.wiley.com/10.1111/gcb.12121> (Accessed May 12, 2018).
- 39 Howden, S. M., F. N. Soussana, Jean-François Tubiello, N. Chhetri, M. and Dunlop, and H. Meinke,
40 2007: Adapting Agriculture to Climate Change. *Proc. Natl. Acad. Sci.*, **104**, 19691–19696,
41 doi:10.1073/pnas.0701890104. <http://www.pnas.org/content/104/50/19691.abstract>.
- 42 Hu, S., Z. Niu, Y. Chen, L. Li, and H. Zhang, 2017: Global wetlands: Potential distribution, wetland
43 loss, and status. *Sci. Total Environ.*, **586**, 319–327, doi:10.1016/j.scitotenv.2017.02.001.
44 <https://linkinghub.elsevier.com/retrieve/pii/S0048969717302425> (Accessed April 16, 2019).
- 45 Hua, F., and Coauthors, 2016: Opportunities for biodiversity gains under the world's largest
46 reforestation programme. *Nat. Commun.*, **7**, doi:10.1038/ncomms12717.
- 47 Huang, G., and Coauthors, 2018: Performance, Economics and Potential Impact of Perennial Rice
48 PR23 Relative to Annual Rice Cultivars at Multiple Locations in Yunnan Province of China.

- 1 *Sustainability*, **10**, 1086, doi:10.3390/su10041086. <http://www.mdpi.com/2071-1050/10/4/1086>
2 (Accessed May 21, 2018).
- 3 Huang, J., T. Wang, W. Wang, Z. Li, and H. Yan, 2014: Climate effects of dust aerosols over East
4 Asian arid and semiarid regions. *J. Geophys. Res. Atmos.*, **119**, 11,398–11,416,
5 doi:10.1002/2014JD021796. <http://doi.wiley.com/10.1002/2014JD021796> (Accessed March 16,
6 2019).
- 7 Huggel, C., J. J. Clague, and O. Korup, 2012: Is climate change responsible for changing landslide
8 activity in high mountains? *Earth Surf. Process. Landforms*, **37**, 77–91, doi:10.1002/esp.2223.
9 <http://doi.wiley.com/10.1002/esp.2223> (Accessed April 10, 2019).
- 10 Hulme, P. E., 2017: Climate change and biological invasions: evidence, expectations, and response
11 options. *Biol. Rev.*, **92**, 1297–1313, doi:10.1111/brv.12282.
12 <http://doi.wiley.com/10.1111/brv.12282> (Accessed March 18, 2019).
- 13 Hungate, B. A., and Coauthors, 2017: The economic value of grassland species for carbon storage.
14 *Sci. Adv.*, **3**, e1601880, doi:10.1126/sciadv.1601880.
15 <http://advances.sciencemag.org/lookup/doi/10.1126/sciadv.1601880> (Accessed May 21, 2018).
- 16 Hussain, S., T. Siddique, M. Saleem, M. Arshad, and A. Khalid, 2009: Chapter 5 Impact of Pesticides
17 on Soil Microbial Diversity, Enzymes, and Biochemical Reactions. *Adv. Agron.*, **102**, 159–200,
18 doi:10.1016/S0065-2113(09)01005-0.
19 https://www.researchgate.net/profile/Muhammad_Saleem22/publication/223688980_Impact_of_Pesticides_on_Soil_Microbial_Diversity_Enzymes_and_Biochemical_Reactions/links/59db7dc4aca272ab722b76ec/Impact-of-Pesticides-on-Soil-Microbial-Diversity-Enzymes-and-Bio
20 21 [aca272ab722b76ec/Impact-of-Pesticides-on-Soil-Microbial-Diversity-Enzymes-and-Bio](https://www.researchgate.net/profile/Muhammad_Saleem22/publication/223688980_Impact_of_Pesticides_on_Soil_Microbial_Diversity_Enzymes_and_Biochemical_Reactions/links/59db7dc4aca272ab722b76ec/Impact-of-Pesticides-on-Soil-Microbial-Diversity-Enzymes-and-Bio)
21 22 (Accessed October 1, 2018).
- 23 HÜVE, K., I. BICHELE, B. RASULOV, and Ü. NIINEMETS, 2011: When it is too hot for
24 photosynthesis: heat-induced instability of photosynthesis in relation to respiratory burst, cell
25 permeability changes and H₂O₂ formation. *Plant. Cell Environ.*, **34**, 113–126,
26 doi:10.1111/j.1365-3040.2010.02229.x. <http://doi.wiley.com/10.1111/j.1365-3040.2010.02229.x>
27 (Accessed October 25, 2018).
- 28 Huxman, T. E., and Coauthors, 2004: Convergence across biomes to a common rain-use efficiency.
29 *Nature*, **429**, 651–654, doi:10.1038/nature02561.
30 <http://www.nature.com/doi/10.1038/nature02561> (Accessed January 3, 2018).
- 31 IEA, 2017: *World Energy Outlook 2017*. Paris,.
- 32 Illius, A., T. O.-E. applications, and undefined 1999, On the relevance of nonequilibrium concepts to
33 arid and semiarid grazing systems. *Wiley Online Libr.*,
34 [http://onlinelibrary.wiley.com/doi/10.1890/1051-0761\(1999\)009\[0798:OTRONC\]2.0.CO;2/full](http://onlinelibrary.wiley.com/doi/10.1890/1051-0761(1999)009[0798:OTRONC]2.0.CO;2/full)
35 (Accessed December 28, 2017).
- 36 Imai, N., T. Seino, S. Aiba, M. Takyu, J. Titin, and K. Kitayama, 2012: Effects of selective logging on
37 tree species diversity and composition of Bornean tropical rain forests at different spatial scales.
38 *Plant Ecol.*, **213**, 1413–1424, doi:10.1007/s11258-012-0100-y.
39 <http://link.springer.com/10.1007/s11258-012-0100-y> (Accessed March 18, 2019).
- 40 Jordan, C. M., X. Hu, A. Arvesen, P. Kauppi, and F. Cherubini, 2018: Contribution of forest wood
41 products to negative emissions: Historical comparative analysis from 1960 to 2015 in Norway,
42 Sweden and Finland. *Carbon Balance Manag.*, **13**, doi:10.1186/s13021-018-0101-9.
- 43 IPCC, 2006: *2006 IPCC Guidelines for National Greenhouse Gas Inventories, Prepared by the*
44 *National Greenhouse Gas Inventories Programme*. Tokyo, Japan, [https://www.ipcc-](https://www.ipcc-nggip.iges.or.jp/public/2006gl/pdf/0_Overview/V0_0_Cover.pdf)
45 [nggip.iges.or.jp/public/2006gl/pdf/0_Overview/V0_0_Cover.pdf](https://www.ipcc-nggip.iges.or.jp/public/2006gl/pdf/0_Overview/V0_0_Cover.pdf).
- 46 —, 2012: *Managing the risks of extreme events and disasters to advance climate change*
47 *adaptation : special report of the Intergovernmental Panel on Climate Change*. Cambridge
48 University Press, Cambridge, UK and New York, USA, 582 pp.

- 1 —, 2013a: Annex I: Atlas of Global and Regional Climate Projections. : *Climate Change 2013: The*
2 *Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the*
3 *Intergovernmental Panel on Climate Change*, 1313–1390.
- 4 —, 2013b: Summary for Policy Makers. *Climate Change 2013: The Physical Science Basis.*
5 *Contribution of Working Group I to the Fifth Assessment Report of the Intergovernmental Panel*
6 *on Climate Change*, V.B. and P.M.M. Stocker, T.F., D. Qin, G.-K. Plattner, M. Tignor, S.K.
7 Allen, J. Boschung, A. Nauels, Y. Xia, Ed., Cambridge University Press, Cambridge, UK and
8 New York, USA, p. 1535.
- 9 —, 2014a: *2013 Supplement to the 2006 Inter-Governmental Panel on Climate Change Guidelines*
10 *for National Greenhouse Gas Inventories: Wetlands. IPCC, Switzerland.*
- 11 —, 2014b: Summary for Policymakers. *Climate Change 2014: Mitigation of Climate Change.*
12 *Contribution of Working Group III to the Fifth Assessment Report of the Intergovernmental*
13 *Panel on Climate Change*, O. Edenhofer et al., Eds., Cambridge University Press, Cambridge,
14 United Kingdom and New York, NY, USA., 1–34.
- 15 —, 2018a: Summary for Policymakers. *Global Warming of 1.5 °C an IPCC special report on the*
16 *impacts of global warming of 1.5 °C above pre-industrial levels and related global greenhouse*
17 *gas emission pathways, in the context of strengthening the global response to the threat of*
18 *climate change* http://report.ipcc.ch/sr15/pdf/sr15_spm_final.pdf (Accessed October 29, 2018).
- 19 —, 2018b: *Global Warming of 1.5 °C an IPCC special report on the impacts of global warming of*
20 *1.5 °C above pre-industrial levels and related global greenhouse gas emission pathways, in the*
21 *context of strengthening the global response to the threat of climate change.*
22 <http://www.ipcc.ch/report/sr15/>.
- 23 IPCC, 2014: Annex II, 2014c: *Glossary [Mach, K.J., S. Planton and C. von Stechow (eds.)]. In:*
24 *Climate Change 2014: Synthesis Report. Contribution of Working Groups I, II and III to the*
25 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change [Core Writing*
26 *Team, R.K. P. 117-130 pp.*
- 27 IPCC WGII, 2014: *Climate Change 2014. Glossary.* 20 pp.
- 28 Islam, M. S., A. T. Wong, M. S. Islam, and A. T. Wong, 2017: Climate Change and Food In/Security:
29 A Critical Nexus. *Environments*, **4**, 38, doi:10.3390/environments4020038.
30 <http://www.mdpi.com/2076-3298/4/2/38> (Accessed March 5, 2019).
- 31 Itkonen, P., 2016: Land rights as the prerequisite for Sámi culture: Skolt Sámi's changing relation to
32 nature in Finland. *Indigenous Rights in Modern Landscapes*, L. Elenius, C. Allard, and C.
33 Sandström, Eds., Routledge, 94–105
34 [https://www.taylorfrancis.com/books/e/9781317059684/chapters/10.4324%2F9781315607559-](https://www.taylorfrancis.com/books/e/9781317059684/chapters/10.4324%2F9781315607559-12)
35 [12](https://www.taylorfrancis.com/books/e/9781317059684/chapters/10.4324%2F9781315607559-12) (Accessed May 16, 2018).
- 36 IUCN, 2003: . *Environmental degradation and impacts on livelihood: Sea intrusion - a case study.*
37 *International Union for Conservation of Nature, Pakistan. Sindh Programme Office.*
38 https://cmsdata.iucn.org/downloads/pk_environmenta pp.
- 39 Jagger, P., K. Lawlor, M. Brockhaus, M. F. Gebara, D. J. Sonwa, and I. A. P. Resosudarmo, 2015:
40 REDD+ safeguards in national policy discourse and pilot projects. *Analysing REDD+:*
41 *Challenges and choices*, L. Angelsen, A., Brockhaus, M., Sunderlin, W.D. Verchot, Ed.,
42 CIFOR, Bogor, Indonesia <https://cgspace.cgiar.org/handle/10568/95003> (Accessed March 26,
43 2019).
- 44 Jastrow, J. D., J. E. Amonette, and V. L. Bailey, 2007: Mechanisms controlling soil carbon turnover
45 and their potential application for enhancing carbon sequestration. *Clim. Change*, **80**, 5–23,
46 doi:10.1007/s10584-006-9178-3. <http://link.springer.com/10.1007/s10584-006-9178-3>
47 (Accessed May 21, 2018).
- 48 Jeffery, S., F. G. A. Verheijen, C. Kammann, and D. Abalos, 2016: Biochar effects on methane
49 emissions from soils: A meta-analysis. *Soil Biol. Biochem.*, **101**, 251–258,

- 1 doi:10.1016/J.SOILBIO.2016.07.021.
2 <https://www.sciencedirect.com/science/article/pii/S0038071716301663> (Accessed December 1,
3 2017).
- 4 Jerneck, A., 2018a: What about Gender in Climate Change? Twelve Feminist Lessons from
5 Development. *Sustainability*, **10**, 627, doi:10.3390/su10030627. [http://www.mdpi.com/2071-](http://www.mdpi.com/2071-1050/10/3/627)
6 [1050/10/3/627](http://www.mdpi.com/2071-1050/10/3/627) (Accessed April 16, 2018).
- 7 —, 2018b: Taking gender seriously in climate change adaptation and sustainability science
8 research: views from feminist debates and sub-Saharan small-scale agriculture. *Sustain. Sci.*, **13**,
9 403–416, doi:10.1007/s11625-017-0464-y. <http://link.springer.com/10.1007/s11625-017-0464-y>
10 (Accessed May 16, 2018).
- 11 —, and L. Olsson, 2013: More than trees! Understanding the agroforestry adoption gap in
12 subsistence agriculture: Insights from narrative walks in Kenya. *J. Rural Stud.*, **32**, 114–125,
13 doi:10.1016/J.JRURSTUD.2013.04.004.
14 <https://www.sciencedirect.com/science/article/pii/S0743016713000302> (Accessed May 24,
15 2018).
- 16 —, and —, 2014: Food first! Theorising assets and actors in agroforestry: risk evaders,
17 opportunity seekers and ‘the food imperative’ in sub-Saharan Africa. *Int. J. Agric. Sustain.*, **12**,
18 1–22, doi:10.1080/14735903.2012.751714.
19 <http://www.tandfonline.com/doi/abs/10.1080/14735903.2012.751714> (Accessed May 24, 2018).
- 20 Ji, C., K. Cheng, D. Nayak, and G. Pan, 2018: Environmental and economic assessment of crop
21 residue competitive utilization for biochar, briquette fuel and combined heat and power
22 generation. *J. Clean. Prod.*, **192**, 916–923, doi:10.1016/J.JCLEPRO.2018.05.026.
23 <https://www.sciencedirect.com/science/article/pii/S0959652618313544> (Accessed April 7,
24 2019).
- 25 Jia, Y., Zhou, Z., Qiu, Y., 2006: The potential effect of water yield reduction caused by land
26 conversion in the Yellow River basin. In: Zhang L, Bennett J, Wang XH, Xie C, and Zhao. A.
27 *Study of Sustainable Use of Land Resources in Northwestern China*. Beijing: China National
28 *Forestry Economics and Development Research Center*.
- 29 Jobbágy, E. G., and O. E. Sala, 2000: CONTROLS OF GRASS AND SHRUB ABOVEGROUND
30 PRODUCTION IN THE PATAGONIAN STEPPE. *Ecol. Appl.*, **10**, 541–549,
31 doi:10.1890/1051-0761(2000)010[0541:COGASA]2.0.CO;2.
32 [http://onlinelibrary.wiley.com/doi/10.1890/1051-](http://onlinelibrary.wiley.com/doi/10.1890/1051-0761(2000)010%5B0541:COGASA%5D2.0.CO;2/full)
33 [0761\(2000\)010%5B0541:COGASA%5D2.0.CO;2/full](http://onlinelibrary.wiley.com/doi/10.1890/1051-0761(2000)010%5B0541:COGASA%5D2.0.CO;2/full) (Accessed January 3, 2018).
- 34 —, T. Tóth, M. D. Noretto, and S. Earman, 2017: On the Fundamental Causes of High
35 Environmental Alkalinity (pH ≥ 9): An Assessment of Its Drivers and Global Distribution. *L.*
36 *Degrad. Dev.*, **28**, 1973–1981, doi:10.1002/ldr.2718. <http://doi.wiley.com/10.1002/ldr.2718>
37 (Accessed March 18, 2019).
- 38 John, R., and Coauthors, 2016: Differentiating anthropogenic modification and precipitation-driven
39 change on vegetation productivity on the Mongolian Plateau. *Landsc. Ecol.*, **31**, 547–566,
40 doi:10.1007/s10980-015-0261-x. <http://link.springer.com/10.1007/s10980-015-0261-x>
41 (Accessed October 28, 2018).
- 42 Johnson, D. L., and L. A. Lewis, 2007: *Land degradation: creation and destruction*. Rowman &
43 Littlefield, 303 pp.
- 44 Johnson, J. M.-F., R. R. Allmaras, and D. C. Reicosky, 2006: Estimating Source Carbon from Crop
45 Residues, Roots and Rhizodeposits Using the National Grain-Yield Database. *Agron. J.*, **98**, 622,
46 doi:10.2134/agronj2005.0179. <https://www.agronomy.org/publications/aj/abstracts/98/3/622>
47 (Accessed May 21, 2018).
- 48 Johnson, J. M., L. J. Moore, K. Ells, A. B. Murray, P. N. Adams, R. A. MacKenzie, and J. M. Jaeger,
49 2015: Recent shifts in coastline change and shoreline stabilization linked to storm climate

- 1 change. *Earth Surf. Process. Landforms*, **40**, 569–585, doi:10.1002/esp.3650.
2 <http://doi.wiley.com/10.1002/esp.3650> (Accessed March 16, 2019).
- 3 Jolly, W. M., M. A. Cochrane, P. H. Freeborn, Z. A. Holden, T. J. Brown, G. J. Williamson, and D.
4 M. J. S. Bowman, 2015: Climate-induced variations in global wildfire danger from 1979 to
5 2013. *Nat. Commun.*, **6**, 7537, doi:10.1038/ncomms8537.
6 <http://www.nature.com/articles/ncomms8537> (Accessed March 16, 2019).
- 7 de Jong, J., C. Akselsson, G. Egnell, S. Löfgren, and B. A. Olsson, 2017: Realizing the energy
8 potential of forest biomass in Sweden – How much is environmentally sustainable? *For. Ecol.*
9 *Manage.*, **383**, 3–16, doi:10.1016/J.FORECO.2016.06.028.
10 <https://www.sciencedirect.com/science/article/pii/S0378112716303255> (Accessed April 11,
11 2019).
- 12 Joosten, H., Tanneberger, F., 2017: Peatland use in Europe. In: Joosten H., Tanneberger F., Moen A.,
13 (Eds.), Mires and peatlands of Europe: Status, distribution and conservation. Schweizerbart
14 Science Publisher. Stuttgart. 151–173.
- 15 Jorgenson, M. T., and T. E. Osterkamp, 2005: Response of boreal ecosystems to varying modes of
16 permafrost degradation. *Can. J. For. Res.*, **35**, 2100–2111, doi:10.1139/x05-153.
17 <http://www.nrcresearchpress.com/doi/10.1139/x05-153> (Accessed May 23, 2018).
- 18 Joseph, J., 2013: Resilience as embedded neoliberalism: a governmentality approach. *Resilience*, **1**,
19 38–52, doi:10.1080/21693293.2013.765741.
20 <http://www.tandfonline.com/doi/abs/10.1080/21693293.2013.765741> (Accessed May 25, 2018).
- 21 Joseph, S., and Coauthors, 2013: Shifting paradigms: development of high-efficiency biochar
22 fertilizers based on nano-structures and soluble components. *Carbon Manag.*, **4**, 323–343,
23 doi:10.4155/cmt.13.23. <http://dx.doi.org/10.4155/cmt.13.23>.
- 24 Joseph, S., and Coauthors, 2018: Microstructural and associated chemical changes during the
25 composting of a high temperature biochar: Mechanisms for nitrate, phosphate and other nutrient
26 retention and release. *Sci. Total Environ.*, doi:10.1016/j.scitotenv.2017.09.200.
- 27 Joseph, S. D., and Coauthors, 2010: An investigation into the reactions of biochar in soil. *Australian*
28 *Journal of Soil Research*.
- 29 JRC, 2018: Land Productivity Dynamics. *World Atlas Desertif.*,
30 <https://wad.jrc.ec.europa.eu/landproductivity>.
- 31 Juday, G. P., C. Alix, and T. A. Grant, 2015: Spatial coherence and change of opposite white spruce
32 temperature sensitivities on floodplains in Alaska confirms early-stage boreal biome shift. *For.*
33 *Ecol. Manage.*, **350**, 46–61, doi:10.1016/J.FORECO.2015.04.016.
34 <https://www.sciencedirect.com/science/article/pii/S0378112715002212> (Accessed March 14,
35 2019).
- 36 Kalhoro, N. ., Z. He, D. Xu, M. Faiz, L. V. Yafei, N. Sohoo, and B. A.H., 2016: Vulnerability of the
37 Indus River delta of the North Arabian sea, Pakistan. *Glob. Nest J. (gnest)*. (**18**, 599–610).
- 38 Kalhoro, N. A., Z. He, D. Xu, I. Muhammad, and A. F. Sohoo, 2017: Seasonal variation of
39 oceanographic processes in Indus river estuary. *Seas. Var. Oceanogr. Process. Indus river*
40 *estuary.MAUSAM*, **68**, 643–654.
- 41 KAMMANN, C., and Coauthors, 2017: BIOCHAR AS A TOOL TO REDUCE THE
42 AGRICULTURAL GREENHOUSE-GAS BURDEN – KNOWN, UNKNOWN AND
43 FUTURE RESEARCH NEEDS. *J. Environ. Eng. Landsc. Manag.*, **25**, 114–139,
44 doi:10.3846/16486897.2017.1319375.
45 <http://journals.vgtu.lt/index.php/JEELM/article/view/1632> (Accessed April 2, 2019).
- 46 Kaplan, J. O., K. M. Krumhardt, E. C. Ellis, W. F. Ruddiman, C. Lemmen, and K. K. Goldewijk,
47 2011: Holocene carbon emissions as a result of anthropogenic land cover change. *The Holocene*,
48 **21**, 775–791, doi:10.1177/0959683610386983.

- 1 <http://journals.sagepub.com/doi/10.1177/0959683610386983> (Accessed May 12, 2018).
- 2 Karami, M., M. Afyuni, A. H. Khoshgoftarmanesh, A. Papritz, and R. Schulin, 2009: Grain Zinc,
3 Iron, and Copper Concentrations of Wheat Grown in Central Iran and Their Relationships with
4 Soil and Climate Variables. *J. Agric. Food Chem.*, **57**, 10876–10882, doi:10.1021/jf902074f.
5 <http://pubs.acs.org/doi/abs/10.1021/jf902074f> (Accessed May 31, 2018).
- 6 Karbassi, A., G. N. Bidhendi, A. Pejman, and M. E. Bidhendi, 2010: Environmental impacts of
7 desalination on the ecology of Lake Urmia. *J. Great Lakes Res.*, **36**, 419–424,
8 doi:10.1016/j.jglr.2010.06.004.
- 9 Kassa, H., S. Dondeyne, J. Poesen, A. Frankl, and J. Nyssen, 2017: Transition from Forest-based to
10 Cereal-based Agricultural Systems: A Review of the Drivers of Land use Change and
11 Degradation in Southwest Ethiopia. *L. Degrad. Dev.*, **28**, 431–449, doi:10.1002/ldr.2575.
12 <http://doi.wiley.com/10.1002/ldr.2575> (Accessed May 16, 2018).
- 13 Kassam, A., and Coauthors, 2013: Sustainable soil management is more than what and how crops are
14 grown. *Principles of Sustainable Soil Management in Agroecosystems*, R. Lal and B.A. Stewart,
15 Eds., CRC Press, Boca Raton, FL, USA, 337–400
16 <https://www.taylorfrancis.com/books/9781466513471> (Accessed November 2, 2018).
- 17 Kates, R. W., W. R. Travis, and T. J. Wilbanks, 2012: Transformational adaptation when incremental
18 adaptations to climate change are insufficient. *Proc. Natl. Acad. Sci. U. S. A.*, **109**, 7156–7161,
19 doi:10.1073/pnas.1115521109. <http://www.ncbi.nlm.nih.gov/pubmed/22509036> (Accessed
20 October 19, 2018).
- 21 Kauppi, P. E., M. Posch, and P. Pirinen, 2014: Large impacts of climatic warming on growth of boreal
22 forests since 1960. *PLoS One*, **9**, 1–6, doi:10.1371/journal.pone.0111340.
- 23 Kauppi, P. E., V. Sandström, and A. Lipponen, 2018: Forest resources of nations in relation to human
24 well-being. *PLoS One*, **13**, e0196248, doi:10.1371/journal.pone.0196248.
25 <https://doi.org/10.1371/journal.pone.0196248>.
- 26 Kaye, J. P., and M. Quemada, 2017: Using cover crops to mitigate and adapt to climate change. A
27 review. *Agron. Sustain. Dev.*, **37**, 4, doi:10.1007/s13593-016-0410-x.
28 <http://link.springer.com/10.1007/s13593-016-0410-x> (Accessed March 28, 2019).
- 29 Keenan, R. J., G. A. Reams, F. Achard, J. V. de Freitas, A. Grainger, and E. Lindquist, 2015:
30 Dynamics of global forest area: Results from the FAO Global Forest Resources Assessment
31 2015. *For. Ecol. Manage.*, **352**, 9–20, doi:10.1016/J.FORECO.2015.06.014.
32 <https://www.sciencedirect.com/science/article/pii/S0378112715003400> (Accessed October 31,
33 2018).
- 34 Keenan, T. F., D. Y. Hollinger, G. Bohrer, D. Dragoni, J. W. Munger, H. P. Schmid, and A. D.
35 Richardson, 2013: Increase in forest water-use efficiency as atmospheric carbon dioxide
36 concentrations rise. *Nature*, **499**, 324–327, doi:10.1038/nature12291.
37 <http://www.nature.com/articles/nature12291> (Accessed March 5, 2018).
- 38 —, I. C. Prentice, J. G. Canadell, C. A. Williams, H. Wang, M. Raupach, and G. J. Collatz, 2017:
39 Corrigendum: Recent pause in the growth rate of atmospheric CO₂ due to enhanced terrestrial
40 carbon uptake. *Nat. Commun.*, **8**, 16137, doi:10.1038/ncomms16137.
41 <http://www.nature.com/doi/10.1038/ncomms16137>.
- 42 Keith, H., D. Lindenmayer, B. Mackey, D. Blair, L. Carter, L. McBurney, S. Okada, and T. Konishi-
43 Nagano, 2014: Managing temperate forests for carbon storage: impacts of logging versus forest
44 protection on carbon stocks. *Ecosphere*, **5**, art75, doi:10.1890/ES14-00051.1.
- 45 Kelly (Letcher), R. A., and Coauthors, 2013: Selecting among five common modelling approaches for
46 integrated environmental assessment and management. *Environ. Model. Softw.*, **47**, 159–181,
47 doi:10.1016/J.ENVSFT.2013.05.005.
48 <https://www.sciencedirect.com/science/article/pii/S1364815213001151?via%3Dihub> (Accessed
49 November 2, 2018).

- 1 Kendon, E. J., N. M. Roberts, H. J. Fowler, M. J. Roberts, S. C. Chan, and C. A. Senior, 2014:
2 Heavier summer downpours with climate change revealed by weather forecast resolution model.
3 *Nat. Clim. Chang.*, **4**, 570–576, doi:10.1038/nclimate2258.
4 <http://www.nature.com/doi/10.1038/nclimate2258> (Accessed November 14, 2017).
- 5 Keogh, M. E., and T. E. Törnqvist, 2019: Measuring rates of present-day relative sea-level rise in low-
6 elevation coastal zones: a critical evaluation. *Ocean Sci.*, **15**, 61–73, doi:10.5194/os-15-61-2019.
7 <https://www.ocean-sci.net/15/61/2019/> (Accessed April 7, 2019).
- 8 Kerns, B. K., J. B. Kim, J. D. Kline, and M. A. Day, 2016: US exposure to multiple landscape
9 stressors and climate change. *Reg. Environ. Chang.*, **16**, 2129–2140, doi:10.1007/s10113-016-
10 0934-2. <http://link.springer.com/10.1007/s10113-016-0934-2> (Accessed March 6, 2019).
- 11 Kerr, J. M., J. V. DePinto, D. McGrath, S. P. Sowa, and S. M. Swinton, 2016: Sustainable
12 management of Great Lakes watersheds dominated by agricultural land use. *J. Great Lakes Res.*,
13 **42**, 1252–1259, doi:10.1016/J.JGLR.2016.10.001.
14 <https://www.sciencedirect.com/science/article/pii/S038013301630185X> (Accessed April 17,
15 2019).
- 16 Kessler, C. A., and L. Stroosnijder, 2006: Land degradation assessment by farmers in Bolivian
17 mountain valleys. *L. Degrad. Dev.*, **17**, 235–248, doi:10.1002/ldr.699.
18 <http://doi.wiley.com/10.1002/ldr.699> (Accessed April 28, 2018).
- 19 Keys, P. W., L. Wang-Erlandsson, L. J. Gordon, V. Galaz, and J. Ebbesson, 2017: Approaching
20 moisture recycling governance. *Glob. Environ. Chang.*, **45**, 15–23,
21 doi:10.1016/J.GLOENVCHA.2017.04.007.
22 <https://www.sciencedirect.com/science/article/pii/S0959378016306902> (Accessed March 13,
23 2019).
- 24 Kiage, L. M., 2013: Perspectives on the assumed causes of land degradation in the rangelands of Sub-
25 Saharan Africa. *Prog. Phys. Geogr.*, **37**, 664–684, doi:10.1177/0309133313492543.
- 26 Kidane, D., and B. Alemu, 2015: The Effect of Upstream Land Use Practices on Soil Erosion and
27 Sedimentation in the Upper Blue Nile Basin, Ethiopia. *Res. J. Agric. Environ. Manag.*, **4**, 55–68.
28 [https://www.researchgate.net/profile/Binyam_Alemu/publication/324845331_The_Effect_of_U
29 pstream_Land_Use_Practices_on_Soil_Erosion_and_Sedimentation_in_the_Upper_Blue_Nile_
30 Basin_Ethiopia/links/5ae79275aca2725dabb33f77/The-Effect-of-Upstream-Land-Use-Practi](https://www.researchgate.net/profile/Binyam_Alemu/publication/324845331_The_Effect_of_Upstream_Land_Use_Practices_on_Soil_Erosion_and_Sedimentation_in_the_Upper_Blue_Nile_Basin_Ethiopia/links/5ae79275aca2725dabb33f77/The-Effect-of-Upstream-Land-Use-Practi).
- 31 Kilpeläinen, A., H. Strandman, T. Grönholm, V. P. Ikonen, P. Torssonen, S. Kellomäki, and H.
32 Peltola, 2017: Effects of Initial Age Structure of Managed Norway Spruce Forest Area on Net
33 Climate Impact of Using Forest Biomass for Energy. *Bioenergy Res.*, **10**, 499–508,
34 doi:10.1007/s12155-017-9821-z.
- 35 Kim, G. S., C.-H. H. Lim, S. J. Kim, J. Lee, Y. Y. Y. Son, and W.-K. K. Lee, 2017: Effect of
36 National-Scale Afforestation on Forest Water Supply and Soil Loss in South Korea, 1971–2010.
37 *Sustainability*, **9**, 1017, doi:10.3390/su9061017. <http://www.mdpi.com/2071-1050/9/6/1017>
38 (Accessed May 21, 2018).
- 39 Kim, K. H., and L. Zsuffa, 1994: Reforestation of South Korea: The history and analysis of a unique
40 case in forest tree improvement and forestry. *For. Chron.*, **70**, 58–64, doi:10.5558/tfc70058-1.
41 <http://pubs.cif-ifc.org/doi/pdf/10.5558/tfc70058-1> (Accessed May 21, 2018).
- 42 Kim, M., W. Lee, W. Kurz, D. Kwak, S. Morken, C. Smyth, and D. Ryu, 2016: Estimating carbon
43 dynamics in forest carbon pools under IPCC standards in South Korea using CBM-CFS3.
44 *iForest - Biogeosciences For.*, **10**, 83–92, doi:10.3832/ifor2040-009.
45 <http://www.sisef.it/iforest/?doi=ifor2040-009> (Accessed May 21, 2018).
- 46 Kiruki, H. M., E. H. van der Zanden, Ž. Malek, and P. H. Verburg, 2017: Land Cover Change and
47 Woodland Degradation in a Charcoal Producing Semi-Arid Area in Kenya. *L. Degrad. Dev.*, **28**,
48 472–481, doi:10.1002/ldr.2545. <https://doi.org/10.1002/ldr.2545>.
- 49 Kirwan, M. L., A. B. Murray, J. P. Donnelly, and D. R. Corbett, 2011: Rapid wetland expansion

- 1 during European settlement and its implication for marsh survival under modern sediment
2 delivery rates. *Geology*, **39**, 507–510, doi:10.1130/G31789.1.
3 [http://pubs.geoscienceworld.org/geology/article/39/5/507/130567/Rapid-wetland-expansion-](http://pubs.geoscienceworld.org/geology/article/39/5/507/130567/Rapid-wetland-expansion-during-European-settlement)
4 [during-European-settlement](http://pubs.geoscienceworld.org/geology/article/39/5/507/130567/Rapid-wetland-expansion-during-European-settlement) (Accessed December 27, 2017).
- 5 Klein, R.J.T., G.F. Midgley, B.L. Preston, M. Alam, F.G.H. Berkhout, K. Dow, and M. R. S., 2014:
6 Adaptation opportunities, constraints, and limits. *Climate Change 2014: Impacts, Adaptation,*
7 *and Vulnerability. Part A: Global and Sectoral Aspects. Contribution of Working Group II to the*
8 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, M.D.M. Field,
9 C.B., V.R. Barros, D.J. Dokken, K.J. Mach, S.M. T.E. Bilir, M. Chatterjee, K.L. Ebi, Y.O.
10 Estrada, R.C. Genova, B. Girma, E.S. Kissel, A.N. Levy, and And L.L.W. P.R. Mastrandrea,
11 Eds., Cambridge University Press, Cambridge, UK and New York, USA, 899–943.
- 12 Klemm, W., B. G. Heusinkveld, S. Lenzholzer, and B. van Hove, 2015: Street greenery and its
13 physical and psychological impact on thermal comfort. *Landsc. Urban Plan.*, **138**, 87–98,
14 doi:10.1016/J.LANDURBPLAN.2015.02.009.
15 <https://www.sciencedirect.com/science/article/pii/S0169204615000407> (Accessed March 10,
16 2019).
- 17 Klöck, C., and P. D. Nunn, 2019: Adaptation to Climate Change in Small Island Developing States: A
18 Systematic Literature Review of Academic Research. *J. Environ. Dev.*, 107049651983589,
19 doi:10.1177/1070496519835895. <http://journals.sagepub.com/doi/10.1177/1070496519835895>
20 (Accessed April 10, 2019).
- 21 Knapp, A. K., and M. D. Smith, 2001: Variation Among Biomes in Temporal Dynamics of
22 Aboveground Primary Production. *Science (80-.)*, **291**, 481–484,
23 doi:10.1126/science.291.5503.481.
24 <http://www.sciencemag.org/cgi/doi/10.1126/science.291.5503.481> (Accessed January 3, 2018).
- 25 Knorr, W., L. Jiang, and A. Arneeth, 2016: Climate, CO 2 and human population impacts on global
26 wildfire emissions. *Biogeosciences*, **13**, 267–282, doi:10.5194/bg-13-267-2016.
27 www.biogeosciences.net/13/267/2016/ (Accessed March 16, 2019).
- 28 Knutson, T. R., and Coauthors, 2010: Tropical cyclones and climate change. *Nat. Geosci.*, **3**, 157–
29 163, doi:10.1038/ngeo779. <http://www.nature.com/articles/ngeo779> (Accessed May 23, 2018).
- 30 —, J. J. Sirutis, M. Zhao, R. E. Tuleya, M. Bender, G. A. Vecchi, G. Villarini, and D. Chavas,
31 2015: Global projections of intense tropical cyclone activity for the late twenty-first century
32 from dynamical downscaling of CMIP5/RCP4.5 scenarios. *J. Clim.*, **28**, 7203–7224,
33 doi:10.1175/JCLI-D-15-0129.1.
- 34 Köhl, M., P. R. Neupane, and N. Lotfiomran, 2017: The impact of tree age on biomass growth and
35 carbon accumulation capacity: A retrospective analysis using tree ring data of three tropical tree
36 species grown in natural forests of Suriname. *PLoS One*, **12**, e0181187,
37 doi:10.1371/journal.pone.0181187. <https://dx.plos.org/10.1371/journal.pone.0181187> (Accessed
38 March 26, 2019).
- 39 Kolinjivadi, V. K., and T. Sunderland, 2012: A review of two payment schemes for watershed
40 services from China and Vietnam: The interface of government control and PES theory. *Ecol.*
41 *Soc.*, doi:10.5751/ES-05057-170410.
- 42 Kolton, M., E. R. Graber, L. Tsehansky, Y. Elad, and E. Cytryn, 2017: Biochar-stimulated plant
43 performance is strongly linked to microbial diversity and metabolic potential in the rhizosphere.
44 *New Phytol.*, doi:10.1111/nph.14253.
- 45 Koohafkann, P., and M. A. Altieri, 2011: *Agricultural Heritage Systems: A legacy for the future.*
46 Rome, Italy, 49 pp. [http://worldagriculturalheritage.org/wp-](http://worldagriculturalheritage.org/wp-content/uploads/2014/12/GIAHS_Booklet_EN_WEB2011.pdf)
47 [content/uploads/2014/12/GIAHS_Booklet_EN_WEB2011.pdf](http://worldagriculturalheritage.org/wp-content/uploads/2014/12/GIAHS_Booklet_EN_WEB2011.pdf).
- 48 Koplitz, S. N., and Coauthors, 2016: Public health impacts of the severe haze in Equatorial Asia in
49 September-October 2015: Demonstration of a new framework for informing fire management

- 1 strategies to reduce downwind smoke exposure. *Environ. Res. Lett.*, **11**, doi:10.1088/1748-
2 9326/11/9/094023.
- 3 van de Koppel, J., M. Rietkerk, and F. J. Weissing, 1997: Catastrophic vegetation shifts and soil
4 degradation in terrestrial grazing systems. *Trends Ecol. Evol.*, **12**, 352–356, doi:10.1016/S0169-
5 5347(97)01133-6.
6 <https://www.sciencedirect.com/science/article/pii/S0169534797011336?via%3Dihub> (Accessed
7 May 8, 2018).
- 8 Korea Forest Service, 2017: *Statistical yearbook of forestry*. Korea Forest Service, Ed. Deajeon,
9 Korea, [http://www.scb.se/en/finding-statistics/statistics-by-subject-area/agriculture-forestry-and-
10 fishery/general-statistics/statistical-yearbook-of-forestry/](http://www.scb.se/en/finding-statistics/statistics-by-subject-area/agriculture-forestry-and-fishery/general-statistics/statistical-yearbook-of-forestry/) (Accessed May 21, 2018).
- 11 Korhonen-Kurki, K., and Coauthors, 2018: What drives policy change for REDD+? A qualitative
12 comparative analysis of the interplay between institutional and policy arena factors. *Clim.*
13 *Policy*, **1**–14, doi:10.1080/14693062.2018.1507897.
14 <https://www.tandfonline.com/doi/full/10.1080/14693062.2018.1507897> (Accessed October 31,
15 2018).
- 16 Kosec, K., H. Ghebru, B. Holtemeyer, V. Mueller, and E. Schmidt, 2018: The Effect of Land Access
17 on Youth Employment and Migration Decisions: Evidence from Rural Ethiopia. *Am. J. Agric.*
18 *Econ.*, **100**, 931–954, doi:10.1093/ajae/aax087.
19 <https://academic.oup.com/ajae/article/100/3/931/4781351> (Accessed April 6, 2019).
- 20 Kosmas, C., and Coauthors, 2014: Evaluation and Selection of Indicators for Land Degradation and
21 Desertification Monitoring: Methodological Approach. *Environ. Manage.*, **54**, 951–970,
22 doi:10.1007/s00267-013-0109-6. <https://doi.org/10.1007/s00267-013-0109-6>.
- 23 Kossin, J. P., 2018: A global slowdown of tropical-cyclone translation speed. *Nature*, **558**, 104–107,
24 doi:10.1038/s41586-018-0158-3. <http://www.nature.com/articles/s41586-018-0158-3> (Accessed
25 October 14, 2018).
- 26 Krause, A., and Coauthors, 2018: Large uncertainty in carbon uptake potential of land-based climate-
27 change mitigation efforts. *Glob. Chang. Biol.*, **24**, 3025–3038, doi:10.1111/gcb.14144.
28 <http://doi.wiley.com/10.1111/gcb.14144> (Accessed April 12, 2019).
- 29 Kreamer, D. K., 2012: The Past, Present, and Future of Water Conflict and International Security. *J.*
30 *Contemp. Water Res. Educ.*, **149**, 87–95, doi:10.1111/j.1936-704X.2012.03130.x.
31 <http://doi.wiley.com/10.1111/j.1936-704X.2012.03130.x> (Accessed January 9, 2018).
- 32 Kristjanson, P., and Coauthors, 2017: Addressing gender in agricultural research for development in
33 the face of a changing climate: where are we and where should we be going? *Int. J. Agric.*
34 *Sustain.*, **15**, 482–500, doi:10.1080/14735903.2017.1336411.
35 <https://www.tandfonline.com/doi/full/10.1080/14735903.2017.1336411> (Accessed October 18,
36 2018).
- 37 Kroeger, A., S. Koenig, A. Thomson, C. Streck, P.-H. Weiner, and H. Bakhtary, 2017: *Forest- and*
38 *Climate-Smart Cocoa in Côte d’Ivoire and Ghana, Aligning Stakeholders to Support*
39 *Smallholders in Deforestation-Free Cocoa*. World Bank, Washington, DC,
40 <https://elibrary.worldbank.org/doi/abs/10.1596/29014>.
- 41 Kulakowski, D., T. T. Veblen, and P. Bebi, 2003: Effects of fire and spruce beetle outbreak legacies
42 on the disturbance regime of a subalpine forest in Colorado. *J. Biogeogr.*, **30**, 1445–1456,
43 doi:10.1046/j.1365-2699.2003.00912.x. <http://doi.wiley.com/10.1046/j.1365-2699.2003.00912.x>
44 (Accessed May 8, 2018).
- 45 Kumar, N., and A. R. Quisumbing, 2015: Policy Reform toward Gender Equality in Ethiopia: Little
46 by Little the Egg Begins to Walk. *World Dev.*, **67**, 406–423,
47 doi:10.1016/J.WORLDDEV.2014.10.029.
48 <https://www.sciencedirect.com/science/article/pii/S0305750X14003441> (Accessed October 18,
49 2018).

- 1 Kuppusamy, S., P. Thavamani, M. Megharaj, K. Venkateswarlu, and R. Naidu, 2016: Agronomic and
2 remedial benefits and risks of applying biochar to soil: Current knowledge and future research
3 directions. *Environ. Int.*, **87**, 1–12, doi:10.1016/J.ENVINT.2015.10.018.
4 <https://www.sciencedirect.com/science/article/pii/S0160412015300842> (Accessed April 2,
5 2019).
- 6 Kurz, W. A., S. J. Beukema, and M. J. Apps, 1998a: Carbon budget implications of the transition fro
7 m natural to managed disturbance regimes in forest landscapes. *Mitig. Adapt. Strateg. Glob.
8 Chang.*, **2**, 405–421.
- 9 Kurz, W. A., S. J. Beukema, and M. J. Apps, 1998b: Carbon Budget Implications of the Transition
10 from Natural to Managed Disturbance Regimes in Forest Landscapes. *Mitig. Adapt. Strateg.
11 Glob. Chang.*, **2**, 405–421, doi:10.1023/b:miti.0000004486.62808.29.
- 12 Kurz, W. A., C. C. Dymond, G. Stinson, G. J. Rampley, E. T. Neilson, A. L. Carroll, T. Ebata, and L.
13 Safranyik, 2008: Mountain pine beetle and forest carbon feedback to climate change. *Nature*,
14 **452**, 987–990, doi:10.1038/nature06777. <http://www.nature.com/articles/nature06777> (Accessed
15 May 12, 2018).
- 16 Kurz, W. A., and Coauthors, 2013: Carbon in Canada’s boreal forest — A synthesis. *Environ. Rev.*,
17 **21**, 260–292, doi:10.1139/er-2013-0041. [http://www.nrcresearchpress.com/doi/10.1139/er-2013-
18 0041](http://www.nrcresearchpress.com/doi/10.1139/er-2013-0041) (Accessed October 20, 2017).
- 19 —, C. Smyth, and T. Lemprière, 2016: Climate change mitigation through forest sector activities:
20 principles, potential and priorities. *Unasylva*, **67**, 61–67.
- 21 Van der Laan, C., B. Wicke, P. A. Verweij, and A. P. C. Faaij, 2017: Mitigation of unwanted direct
22 and indirect land-use change - an integrated approach illustrated for palm oil, pulpwood, rubber
23 and rice production in North and East Kalimantan, Indonesia. *GCB Bioenergy*, **9**, 429–444,
24 doi:10.1111/gcbb.12353. <http://doi.wiley.com/10.1111/gcbb.12353> (Accessed March 14, 2019).
- 25 Labrière, N., B. Locatelli, Y. Laumonier, V. Freycon, and M. Bernoux, 2015: Soil erosion in the
26 humid tropics: A systematic quantitative review. *Agric. Ecosyst. Environ.*, **203**, 127–139,
27 doi:10.1016/J.AGEE.2015.01.027.
28 <https://www.sciencedirect.com/science/article/pii/S0167880915000468> (Accessed February 20,
29 2019).
- 30 Laderach, P., M. Lundy, A. Jarvis, J. Ramirez, E. P. Portilla, K. Schepp, and A. Eitzinger, 2011:
31 Predicted Impact of Climate Change on Coffee Supply Chains. Springer, Berlin, Heidelberg,
32 703–723 http://link.springer.com/10.1007/978-3-642-14776-0_42 (Accessed May 23, 2018).
- 33 Ladha, J. K., H. Pathak, T. J. Krupnik, J. Six, and C. van Kessel, 2005: Efficiency of Fertilizer
34 Nitrogen in Cereal Production: Retrospects and Prospects. *Adv. Agron.*, **87**, 85–156,
35 doi:10.1016/S0065-2113(05)87003-8.
36 <https://www.sciencedirect.com/science/article/pii/S0065211305870038> (Accessed October 18,
37 2018).
- 38 Lado, M., M. Ben-Hur, and I. Shainberg, 2004: Soil Wetting and Texture Effects on Aggregate
39 Stability, Seal Formation, and Erosion. *Soil Sci. Soc. Am. J.*, **68**, 1992,
40 doi:10.2136/sssaj2004.1992. <https://www.soils.org/publications/sssaj/abstracts/68/6/1992>
41 (Accessed November 14, 2017).
- 42 Laganière, J., D. Paré, E. Thiffault, and P. Y. Bernier, 2017: Range and uncertainties in estimating
43 delays in greenhouse gas mitigation potential of forest bioenergy sourced from Canadian forests.
44 *GCB Bioenergy*, **9**, 358–369, doi:10.1111/gcbb.12327.
- 45 Lal, R., 2003: Soil erosion and the global carbon budget. *Environ. Int.*, **29**, 437–450,
46 doi:10.1016/S0160-4120(02)00192-7.
47 <http://linkinghub.elsevier.com/retrieve/pii/S0160412002001927> (Accessed May 7, 2018).
- 48 Lamb, D., 2014: *Large-scale Forest Restoration*. Routledge, London, UK,
49 <https://www.taylorfrancis.com/books/9780203071649> (Accessed May 21, 2018).

- 1 Lambin, E. F., and P. Meyfroidt, 2011: Global land use change, economic globalization, and the
2 looming land scarcity. *Proc. Natl. Acad. Sci. U. S. A.*, **108**, 3465–3472,
3 doi:10.1073/pnas.1100480108. <http://www.ncbi.nlm.nih.gov/pubmed/21321211> (Accessed May
4 25, 2018).
- 5 Lambin, E. F., and Coauthors, 2001: The causes of land-use and land-cover change: moving beyond
6 the myths. *Glob. Environ. Chang.*, **11**, 261–269, doi:10.1016/S0959-3780(01)00007-3.
7 <http://www.sciencedirect.com/science/article/pii/S0959378001000073> (Accessed October 19,
8 2017).
- 9 —, H. J. Geist, and E. Lepers, 2003: DYNAMICS OF LAND-USE AND LAND-COVER C
10 HANGE IN TROPICAL REGIONS. *Annu. Rev. Environ. Resour.*, **28**, 205–241,
11 doi:10.1146/annurev.energy.28.050302.105459.
12 <http://www.annualreviews.org/doi/10.1146/annurev.energy.28.050302.105459> (Accessed
13 October 24, 2018).
- 14 —, and Coauthors, 2018: The role of supply-chain initiatives in reducing deforestation. *Nat. Clim.*
15 *Chang.*, **8**, 109–116, doi:10.1038/s41558-017-0061-1. [http://www.nature.com/articles/s41558-](http://www.nature.com/articles/s41558-017-0061-1)
16 [017-0061-1](http://www.nature.com/articles/s41558-017-0061-1) (Accessed March 29, 2019).
- 17 Lamers, L. P. M., and Coauthors, 2015: Ecological restoration of rich fens in Europe and North
18 America: from trial and error to an evidence-based approach. *Biol. Rev. Camb. Philos. Soc.*,
19 doi:10.1111/brv.12102.
- 20 Lamichhane, J. R., and Coauthors, 2015: Robust cropping systems to tackle pests under climate
21 change. A review. *Agron. Sustain. Dev.*, **35**, 443–459, doi:10.1007/s13593-014-0275-9.
22 <http://link.springer.com/10.1007/s13593-014-0275-9> (Accessed October 24, 2018).
- 23 Landauer, M., and S. Juhola, 2019: Loss and Damage in the Rapidly Changing Arctic. Springer,
24 Cham, 425–447 http://link.springer.com/10.1007/978-3-319-72026-5_18 (Accessed March 30,
25 2019).
- 26 Laniak, G. F., and Coauthors, 2013: Integrated environmental modeling: A vision and roadmap for the
27 future. *Environ. Model. Softw.*, doi:10.1016/j.envsoft.2012.09.006.
- 28 Lau, K. M., and K. M. Kim, 2007: Cooling of the Atlantic by Saharan dust. *Geophys. Res. Lett.*, **34**,
29 n/a-n/a, doi:10.1029/2007GL031538. <http://doi.wiley.com/10.1029/2007GL031538> (Accessed
30 March 16, 2019).
- 31 Lavers, T., 2017: Land Registration and Gender Equality in Ethiopia: How State-Society Relations
32 Influence the Enforcement of Institutional Change. *J. Agrar. Chang.*, **17**, 188–207,
33 doi:10.1111/joac.12138. <http://doi.wiley.com/10.1111/joac.12138> (Accessed October 18, 2018).
- 34 Law, B. E., T. W. Hudiburg, L. T. Berner, J. J. Kent, P. C. Buotte, and M. E. Harmon, 2018: Land use
35 strategies to mitigate climate change in carbon dense temperate forests. *Proc. Natl. Acad. Sci. U.*
36 *S. A.*, **115**, 3663–3668, doi:10.1073/pnas.1720064115.
37 <http://www.ncbi.nlm.nih.gov/pubmed/29555758> (Accessed March 26, 2019).
- 38 Lawhon, M., and J. T. Murphy, 2012: Socio-technical regimes and sustainability transitions. *Prog.*
39 *Hum. Geogr.*, **36**, 354–378, doi:10.1177/0309132511427960.
40 <http://journals.sagepub.com/doi/10.1177/0309132511427960> (Accessed October 31, 2018).
- 41 Lawrence, D., and K. Vandecar, 2015: Effects of tropical deforestation on climate and agriculture.
42 *Nat. Clim. Chang.*, **5**, 27–36, doi:10.1038/nclimate2430.
43 <http://www.nature.com/articles/nclimate2430> (Accessed March 16, 2019).
- 44 Lawson, I. T., and Coauthors, 2015: Improving estimates of tropical peatland area, carbon storage,
45 and greenhouse gas fluxes. *Wetl. Ecol. Manag.*, doi:10.1007/s11273-014-9402-2.
- 46 Le, Q. B., E. Nkonya, and A. Mirzabaev, 2016: Biomass Productivity-Based Mapping of Global Land
47 Degradation Hotspots. *Economics of Land Degradation and Improvement -- A Global*
48 *Assessment for Sustainable Development*, E. Nkonya, A. Mirzabaev, and J. Von Braun, Eds.,

- 1 Springer International Publishing, Cham, 55–84 https://doi.org/10.1007/978-3-319-19168-3_4.
- 2 Lecina-Diaz, J., A. Alvarez, A. Regos, P. Drapeau, A. Paquette, C. Messier, and J. Retana, 2018: The
3 positive carbon stocks-biodiversity relationship in forests: co-occurrence and drivers across five
4 subclimates. *Ecol. Appl.*, **28**, 1481–1493, doi:10.1002/eap.1749.
5 <http://doi.wiley.com/10.1002/eap.1749> (Accessed March 26, 2019).
- 6 Lee, D. K., P. S. Park, and Y. D. Park, 2015: Forest Restoration and Rehabilitation in the Republic of
7 Korea. *Restoration of Boreal and Temperate Forests*, J.A. Stanturf, Ed., CRC Press, Boca
8 Raton, Florida, 230–245
9 <https://www.taylorfrancis.com/books/e/9781482211979/chapters/10.1201%2Fb18809-10>
10 (Accessed May 21, 2018).
- 11 Lee, H.-L., 2009: The impact of climate change on global food supply and demand, food prices, and
12 land use. *Paddy Water Environ.*, **7**, 321–331, doi:10.1007/s10333-009-0181-y.
13 <http://link.springer.com/10.1007/s10333-009-0181-y> (Accessed January 9, 2018).
- 14 Lee, J., and Coauthors, 2014: Estimating the carbon dynamics of South Korean forests from 1954 to
15 2012. *Biogeosciences*, **11**, 4637–4650, doi:10.5194/bg-11-4637-2014.
16 <http://www.biogeosciences.net/11/4637/2014/> (Accessed May 21, 2018).
- 17 Lee, J., C.-H. H. Lim, G. S. Kim, A. Markandya, S. Chowdhury, S. J. Kim, W.-K. K. Lee, and Y. Son,
18 2018a: Economic viability of the national-scale forestation program: The case of success in the
19 Republic of Korea. *Ecosyst. Serv.*, **29**, 40–46, doi:10.1016/j.ecoser.2017.11.001.
20 <https://www.sciencedirect.com/science/article/pii/S2212041617304953> (Accessed May 21,
21 2018).
- 22 Lee, S. G., H.-A. A. Choi, H. Yoo, C. Song, S. Cha, S.-W. W. Bae, Y. Son, and W.-K. K. Lee, 2018b:
23 Restoration plan for degraded forest in the democratic people’s republic of Korea considering
24 suitable tree species and spatial distribution. *Sustain.*, **10**, 856, doi:10.3390/su10030856.
25 <http://www.mdpi.com/2071-1050/10/3/856> (Accessed May 21, 2018).
- 26 Lehmann, J., and M. Kleber, 2015: The contentious nature of soil organic matter. *Nature*, **528**, 60,
27 doi:10.1038/nature16069. <http://www.nature.com/doifinder/10.1038/nature16069> (Accessed
28 March 25, 2019).
- 29 ———, and J. Stephen, 2015: *Biochar for Environmental Management: Science, Technology and*
30 *Implementation*.
- 31 Lei, Y., R. Treuhaft, M. Keller, M. dos-Santos, F. Gonçalves, and M. Neumann, 2018: Quantification
32 of selective logging in tropical forest with spaceborne SAR interferometry. *Remote Sens.*
33 *Environ.*, **211**, 167–183, doi:10.1016/J.RSE.2018.04.009.
34 <https://www.sciencedirect.com/science/article/pii/S0034425718301548> (Accessed March 18,
35 2019).
- 36 Leifeld, J., and L. Menichetti, 2018: The underappreciated potential of peatlands in global climate
37 change mitigation strategies. *Nat. Commun.*, **9**, 1071, doi:10.1038/s41467-018-03406-6.
38 <http://www.nature.com/articles/s41467-018-03406-6> (Accessed March 8, 2019).
- 39 Lemprière, T. C., and Coauthors, 2013: Canadian boreal forests and climate change mitigation.
40 *Environ. Rev.*, **21**, 293–321, doi:10.1139/er-2013-0039.
41 <http://www.nrcresearchpress.com/doi/10.1139/er-2013-0039> (Accessed May 26, 2018).
- 42 Di Leo, N., F. J. Escobedo, and M. Dubbeling, 2016: The role of urban green infrastructure in
43 mitigating land surface temperature in Bobo-Dioulasso, Burkina Faso. *Environ. Dev. Sustain.*,
44 **18**, 373–392, doi:10.1007/s10668-015-9653-y. <http://link.springer.com/10.1007/s10668-015-9653-y>
45 (Accessed May 17, 2018).
- 46 Lesk, C., E. Coffel, A. W. D’Amato, K. Dodds, and R. Horton, 2017: Threats to North American
47 forests from southern pine beetle with warming winters. *Nat. Clim. Chang.*, **7**, 713–717,
48 doi:10.1038/nclimate3375. <http://www.nature.com/doifinder/10.1038/nclimate3375> (Accessed
49 October 22, 2017).

- 1 Lewis, S. L., Y. Malhi, and O. L. Phillips, 2004: Fingerprinting the impacts of global change on
2 tropical forests. *Philos. Trans. R. Soc. Lond. B. Biol. Sci.*, **359**, 437–462,
3 doi:10.1098/rstb.2003.1432. <http://www.ncbi.nlm.nih.gov/pubmed/15212095> (Accessed
4 November 1, 2018).
- 5 Lewis, S. L., C. E. Wheeler, E. T. A. Mitchard, and A. Koch, 2019: Restoring natural forests is the
6 best way to remove atmospheric carbon. *Nature*, **568**, 25–28.
7 <https://www.nature.com/articles/d41586-019-01026-8>.
- 8 Li, D., S. Niu, and Y. Luo, 2012: Global patterns of the dynamics of soil carbon and nitrogen stocks
9 following afforestation: a meta-analysis. *New Phytol.*, **195**, 172–181, doi:10.1111/j.1469-
10 8137.2012.04150.x.
- 11 Li, Q., M. Ma, X. Wu, and H. Yang, 2018a: Snow Cover and Vegetation-Induced Decrease in Global
12 Albedo From 2002 to 2016. *J. Geophys. Res. Atmos.*, **123**, 124–138,
13 doi:10.1002/2017JD027010. <http://doi.wiley.com/10.1002/2017JD027010> (Accessed October 3,
14 2018).
- 15 Li, W., N. MacBean, P. Ciais, P. Defourny, C. Lamarche, S. Bontemps, R. A. Houghton, and S. Peng,
16 2018b: Gross and net land cover changes in the main plant functional types derived from the
17 annual ESA CCI land cover maps (1992–2015). *Earth Syst. Sci. Data*, **10**, 219–234,
18 doi:10.5194/essd-10-219-2018. <https://www.earth-syst-sci-data.net/10/219/2018/> (Accessed
19 February 25, 2019).
- 20 Li, Y., M. Zhao, S. Motesharrei, Q. Mu, E. Kalnay, and S. Li, 2015: Local cooling and warming
21 effects of forests based on satellite observations. *Nat. Commun.*, **6**, doi:10.1038/ncomms7603.
- 22 ———, E. Kalnay, S. Motesharrei, J. Rivas, F. Kucharski, D. Kirk-Davidoff, E. Bach, and N. Zeng,
23 2018c: Climate model shows large-scale wind and solar farms in the Sahara increase rain and
24 vegetation. *Science (80-.)*, **361**, 1019–1022, doi:10.1126/SCIENCE.AAR5629.
25 <http://science.sciencemag.org/content/361/6406/1019> (Accessed March 16, 2019).
- 26 Li, Z., and H. Fang, 2016: Impacts of climate change on water erosion: A review. *Earth-Science Rev.*,
27 **163**, 94–117, doi:10.1016/J.EARSCIREV.2016.10.004.
28 <https://www.sciencedirect.com/science/article/pii/S0012825216303555> (Accessed March 9,
29 2018).
- 30 Liljedahl, A. K., and Coauthors, 2016: Pan-Arctic ice-wedge degradation in warming permafrost and
31 its influence on tundra hydrology. *Nat. Geosci.*, **9**, 312–318, doi:10.1038/ngeo2674.
32 <http://www.nature.com/articles/ngeo2674> (Accessed March 16, 2019).
- 33 Lilleskov, E., and Coauthors, 2019: Is Indonesian peatland loss a cautionary tale for Peru? A two-
34 country comparison of the magnitude and causes of tropical peatland degradation. *Mitig. Adapt.*
35 *Strateg. Glob. Chang.*, **24**, 591–623, doi:10.1007/s11027-018-9790-3.
36 <http://link.springer.com/10.1007/s11027-018-9790-3> (Accessed March 8, 2019).
- 37 Lin, K.-C., S. P. Hamburg, L. Wang, C.-T. Duh, C.-M. Huang, C.-T. Chang, and T.-C. Lin, 2017:
38 Impacts of increasing typhoons on the structure and function of a subtropical forest: reflections
39 of a changing climate. *Sci. Rep.*, **7**, 4911, doi:10.1038/s41598-017-05288-y.
40 <http://www.nature.com/articles/s41598-017-05288-y> (Accessed March 4, 2019).
- 41 Lin, M., and P. Huybers, 2012: Reckoning wheat yield trends. *Environ. Res. Lett.*, **7**, 024016,
42 doi:10.1088/1748-9326/7/2/024016. [http://stacks.iop.org/1748-
43 9326/7/i=2/a=024016?key=crossref.7861e7550adf3d2b7bf0eb905e7c49df](http://stacks.iop.org/1748-9326/7/i=2/a=024016?key=crossref.7861e7550adf3d2b7bf0eb905e7c49df) (Accessed May 23,
44 2018).
- 45 Lin, N., and K. Emanuel, 2016a: Grey swan tropical cyclones. *Nat. Clim. Chang.*, **6**, 106–111,
46 doi:10.1038/nclimate2777. <http://www.nature.com/articles/nclimate2777> (Accessed March 17,
47 2019).
- 48 ———, and ———, 2016b: Grey swan tropical cyclones. *Nat. Clim. Chang.*, **6**, 106–111,
49 doi:10.1038/nclimate2777. <http://www.nature.com/articles/nclimate2777> (Accessed April 10,

- 1 2019).
- 2 Lindenmayer, D. B., C. R. Margules, and D. B. Botkin, 2000: Indicators of Biodiversity for
3 Ecologically Sustainable Forest Management. *Conserv. Biol.*, **14**, 941–950, doi:10.1046/j.1523-
4 1739.2000.98533.x.
- 5 Lindner, M., and Coauthors, 2010: Climate change impacts, adaptive capacity, and vulnerability of
6 European forest ecosystems. *For. Ecol. Manage.*, **259**, 698–709,
7 doi:10.1016/J.FORECO.2009.09.023.
8 <https://www.sciencedirect.com/science/article/pii/S0378112709006604> (Accessed March 5,
9 2018).
- 10 Lindquist, E. J., and R. D’Annunzio, 2016: Assessing global forest land-use change by object-based
11 image analysis. *Remote Sens.*, **8**, doi:10.3390/rs8080678.
- 12 Linham, M. M., and R. J. Nicholls, 2010: Technologies for climate change adaptation: coastal erosion
13 and flooding. *Technol. Clim. Chang. Adapt. Coast. Eros. flooding.*,
14 <https://www.cabdirect.org/cabdirect/abstract/20113337532> (Accessed October 27, 2018).
- 15 Liniger, H. P., M. Studer, R. C. Hauert, and M. Gurtner, 2011: *Sustainable Land Management in*
16 *Practice. Guidelines and Best Practices for Sub-Saharan Africa*. Rome, Italy, 13 pp.
17 <http://www.fao.org/docrep/014/i1861e/i1861e00.pdf>.
- 18 Liu, F., Huang C, He T, Qian X, Liu Y, and L. H., 2002: Role of Grain to Green Program in reducing
19 loss of phosphorus from yellow soil in hilly areas. *J. Soil Water Conserv.*, 20–23.
- 20 Liu, C. L. C., O. Kuchma, and K. V. Krutovsky, 2018a: Mixed-species versus monocultures in
21 plantation forestry: Development, benefits, ecosystem services and perspectives for the future.
22 *Glob. Ecol. Conserv.*, **15**, e00419, doi:10.1016/j.gecco.2018.e00419.
23 <https://doi.org/10.1016/j.gecco.2018.e00419>.
- 24 Liu, J., and J. Diamond, 2008: Science and government: Revolutionizing China’s environmental
25 protection. *Science (80-)*, doi:10.1126/science.1150416.
- 26 Liu, J., S. Li, Z. Ouyang, C. Tam, and X. Chen, 2008: Ecological and socioeconomic effects of
27 China’s policies for ecosystem services. *Proc. Natl. Acad. Sci.*, doi:10.1073/pnas.0706436105.
- 28 Liu, J. Z., W. Ouyang, W. Yang, and S. L. Xu, 2013: in *Encyclopedia of Biodiversity*, S. A. Levin,
29 Academic Press, Waltham, MA, ed. 2. 372–384.
- 30 Liu, S. C., C. Fu, C.-J. Shiu, J.-P. Chen, and F. Wu, 2009: Temperature dependence of global
31 precipitation extremes. *Geophys. Res. Lett.*, **36**, L17702, doi:10.1029/2009GL040218.
32 <http://doi.wiley.com/10.1029/2009GL040218> (Accessed March 14, 2018).
- 33 Liu, T., R. Bruins, and M. Heberling, 2018b: Factors Influencing Farmers’ Adoption of Best
34 Management Practices: A Review and Synthesis. *Sustainability*, **10**, 432,
35 doi:10.3390/su10020432. <http://www.mdpi.com/2071-1050/10/2/432> (Accessed April 16, 2018).
- 36 Liu, Y., and Coauthors, 2018c: Mechanisms of rice straw biochar effects on phosphorus sorption
37 characteristics of acid upland red soils. *Chemosphere*, doi:10.1016/j.chemosphere.2018.05.086.
- 38 Liu, Y. Y., A. I. J. M. van Dijk, R. A. M. de Jeu, J. G. Canadell, M. F. McCabe, J. P. Evans, and G.
39 Wang, 2015a: Recent reversal in loss of global terrestrial biomass. *Nat. Clim. Chang.*, **5**, 470–
40 474, doi:10.1038/nclimate2581. <http://www.nature.com/articles/nclimate2581> (Accessed
41 September 26, 2018).
- 42 —, —, —, —, —, —, and —, 2015b: Recent reversal in loss of global terrestrial
43 biomass. *Nat. Clim. Chang.*, **5**, 470–474, doi:10.1038/nclimate2581.
44 <http://www.nature.com/doi/10.1038/nclimate2581> (Accessed December 20, 2017).
- 45 Long, H. L., G. K. Heilig, J. Wang, X. B. Li, M. Luo, X. Q. Wu, and M. Zhang, 2006: Land use and
46 soil erosion in the upper reaches of the Yangtze River: Some socio-economic considerations on
47 China’s Grain-for-Green Programme. *L. Degrad. Dev.*, doi:10.1002/ldr.736.

- 1 López-Carr, D., 2012: Agro-ecological drivers of rural out-migration to the Maya Biosphere Reserve,
2 Guatemala. *Environ. Res. Lett.*, **7**, 045603, doi:10.1088/1748-9326/7/4/045603.
3 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/7/i=4/a=045603?key=crossref.7030d175e5edb8d437c3072f9d22408e)
4 [9326/7/i=4/a=045603?key=crossref.7030d175e5edb8d437c3072f9d22408e](http://stacks.iop.org/1748-9326/7/i=4/a=045603?key=crossref.7030d175e5edb8d437c3072f9d22408e) (Accessed May 15,
5 2018).
- 6 López-Portillo, J., R. R. Lewis, P. Saenger, A. Rovai, N. Koedam, F. Dahdouh-Guebas, C. Agraz-
7 Hernández, and V. H. Rivera-Monroy, 2017: Mangrove Forest Restoration and Rehabilitation.
8 *Mangrove Ecosystems: A Global Biogeographic Perspective*, Springer International Publishing,
9 Cham, 301–345 http://link.springer.com/10.1007/978-3-319-62206-4_10 (Accessed July 30,
10 2018).
- 11 López-Rosas, H., P. Moreno-Casasola, F. López-Barrera, L. E. Sánchez-Higuero, V. Espejel-
12 González, and J. Vázquez, 2013: Interdune wetland restoration in central Veracruz, Mexico:
13 plant diversity recovery mediated by the hydroperiod. *Restoration of Coastal Dunes*, Springer,
14 Berlin, Heidelberg, 255–269 http://link.springer.com/10.1007/978-3-642-33445-0_16 (Accessed
15 July 18, 2018).
- 16 Lorimer, C. G., and A. S. White, 2003: Scale and frequency of natural disturbances in the northeastern
17 US: Implications for early successional forest habitats and regional age distributions. *For. Ecol.*
18 *Manage.*, **185**, 41–64, doi:10.1016/S0378-1127(03)00245-7.
- 19 Loucks, C., S. Barber-Meyer, M. A. A. Hossain, A. Barlow, and R. M. Chowdhury, 2010: Sea level
20 rise and tigers: predicted impacts to Bangladesh’s Sundarbans mangroves. *Clim. Change*, **98**,
21 291–298, doi:10.1007/s10584-009-9761-5. <http://link.springer.com/10.1007/s10584-009-9761-5>
22 (Accessed June 19, 2018).
- 23 Lovo, S., 2016: Tenure Insecurity and Investment in Soil Conservation. Evidence from Malawi.
24 *World Dev.*, **78**, 219–229, doi:10.1016/J.WORLDDEV.2015.10.023.
25 <https://www.sciencedirect.com/science/article/pii/S0305750X15002454> (Accessed April 3,
26 2019).
- 27 Luetz, J., 2018: Climate Change and Migration in Bangladesh: Empirically Derived Lessons and
28 Opportunities for Policy Makers and Practitioners. *Limits to Climate Change Adaptation*, W.L.
29 Filho and J. Nalau, Eds., Springer International Publishing, Berlin, Heidelberg, 59–105.
- 30 Lukas, M. C., 2014: Eroding battlefields: Land degradation in Java reconsidered. *Geoforum*, **56**, 87–
31 100, doi:10.1016/J.GEOFORUM.2014.06.010.
32 <https://www.sciencedirect.com/science/article/pii/S0016718514001444#b0100> (Accessed May
33 12, 2018).
- 34 Luke, D., K. McLaren, and B. Wilson, 2016: Modeling Hurricane Exposure in a Caribbean Lower
35 Montane Tropical Wet Forest: The Effects of Frequent, Intermediate Disturbances and
36 Topography on Forest Structural Dynamics and Composition. *Ecosystems*, **19**, 1178–1195,
37 doi:10.1007/s10021-016-9993-y. <http://link.springer.com/10.1007/s10021-016-9993-y>
38 (Accessed May 24, 2018).
- 39 Lund, J. F., E. Sungusia, M. B. Mabele, and A. Scheba, 2017: Promising Change, Delivering
40 Continuity: REDD+ as Conservation Fad. *World Dev.*, **89**, 124–139,
41 doi:10.1016/J.WORLDDEV.2016.08.005.
42 <https://www.sciencedirect.com/science/article/pii/S0305750X15312821> (Accessed October 25,
43 2018).
- 44 Lundmark, T., and Coauthors, 2014: Potential roles of Swedish forestry in the context of climate
45 change mitigation. *Forests*, **5**, 557–578, doi:10.3390/f5040557.
- 46 Luo, P., M. Zhou, H. Deng, J. Lyu, W. Cao, K. Takara, D. Nover, and S. Geoffrey Schladow, 2018:
47 Impact of forest maintenance on water shortages: Hydrologic modeling and effects of climate
48 change. *Sci. Total Environ.*, **615**, 1355–1363, doi:10.1016/J.SCITOTENV.2017.09.044.
49 <https://www.sciencedirect.com/science/article/pii/S0048969717323859> (Accessed October 27,

- 1 2018).
- 2 Lutz, A. F., W. W. Immerzeel, A. B. Shrestha, and M. F. P. Bierkens, 2014: Consistent increase in
3 High Asia's runoff due to increasing glacier melt and precipitation. *Nat. Clim. Chang.*, **4**, 587–
4 592, doi:10.1038/nclimate2237. <http://www.nature.com/articles/nclimate2237> (Accessed March
5 6, 2019).
- 6 Luysaert, S., and Coauthors, 2018: Trade-offs in using European forests to meet climate objectives.
7 *Nature*, **562**, 259–262, doi:10.1038/s41586-018-0577-1. <http://dx.doi.org/10.1038/s41586-018-0577-1>.
8
- 9 Lyons-White, J., and A. T. Knight, 2018: Palm oil supply chain complexity impedes implementation
10 of corporate no-deforestation commitments. *Glob. Environ. Chang.*, **50**, 303–313,
11 doi:10.1016/J.GLOENVCHA.2018.04.012.
12 <https://www.sciencedirect.com/science/article/abs/pii/S0959378017310117> (Accessed March
13 29, 2019).
- 14 MA (Millennium Ecosystem Assessment), 2005: *Millennium ecosystem assessment synthesis report*.
15 155 pp.
- 16 Ma, S., T. Zhou, A. Dai, Z. Han, S. Ma, T. Zhou, A. Dai, and Z. Han, 2015: Observed Changes in the
17 Distributions of Daily Precipitation Frequency and Amount over China from 1960 to 2013. *J.*
18 *Clim.*, **28**, 6960–6978, doi:10.1175/JCLI-D-15-0011.1.
19 <http://journals.ametsoc.org/doi/10.1175/JCLI-D-15-0011.1> (Accessed March 7, 2019).
- 20 —, and Coauthors, 2017: Detectable Anthropogenic Shift toward Heavy Precipitation over Eastern
21 China. *J. Clim.*, **30**, 1381–1396, doi:10.1175/JCLI-D-16-0311.1.
22 <http://journals.ametsoc.org/doi/10.1175/JCLI-D-16-0311.1> (Accessed March 7, 2019).
- 23 MacDicken, K. G., P. Sola, J. E. Hall, C. Sabogal, M. Tadoum, and C. de Wasseige, 2015: Global
24 progress toward sustainable forest management. *For. Ecol. Manage.*, **352**, 47–56,
25 doi:10.1016/j.foreco.2015.02.005. <http://dx.doi.org/10.1016/j.foreco.2015.02.005>.
- 26 Macfadyen, S., G. McDonald, and M. P. Hill, 2018: From species distributions to climate change
27 adaptation: Knowledge gaps in managing invertebrate pests in broad-acre grain crops. *Agric.*
28 *Ecosyst. Environ.*, **253**, 208–219, doi:10.1016/J.AGEE.2016.08.029.
29 <https://www.sciencedirect.com/science/article/pii/S0167880916304327> (Accessed March 18,
30 2019).
- 31 Machisa, M., J. Wichmann, and P. S. Nyasulu, 2013: Biomass fuel use for household cooking in
32 Swaziland: is there an association with anaemia and stunting in children aged 6-36 months?
33 *Trans. R. Soc. Trop. Med. Hyg.*, **107**, 535–544, doi:10.1093/trstmh/trt055.
34 <https://academic.oup.com/trstmh/article-lookup/doi/10.1093/trstmh/trt055> (Accessed April 10,
35 2019).
- 36 Mackey, B., I. C. Prentice, W. Steffen, J. I. House, D. Lindenmayer, H. Keith, and S. Berry, 2013:
37 Untangling the confusion around land carbon science and climate change mitigation policy. *Nat.*
38 *Clim. Chang.*, **3**, 552–557, doi:10.1038/nclimate1804.
- 39 —, and Coauthors, 2015: Policy Options for the World's Primary Forests in Multilateral
40 Environmental Agreements. *Conserv. Lett.*, **8**, 139–147, doi:10.1111/conl.12120.
41 <http://doi.wiley.com/10.1111/conl.12120> (Accessed October 19, 2018).
- 42 Maes, J., and S. Jacobs, 2017: Nature-Based Solutions for Europe's Sustainable Development.
43 *Conserv. Lett.*, doi:10.1111/conl.12216.
- 44 Maestre, F. T., and Coauthors, 2009: Shrub encroachment can reverse desertification in semi-arid
45 Mediterranean grasslands. *Ecol. Lett.*, **12**, 930–941, doi:10.1111/j.1461-0248.2009.01352.x.
46 <http://doi.wiley.com/10.1111/j.1461-0248.2009.01352.x> (Accessed May 8, 2018).
- 47 —, and Coauthors, 2013: Changes in biocrust cover drive carbon cycle responses to climate change
48 in drylands. *Glob. Chang. Biol.*, **19**, 3835–3847, doi:10.1111/gcb.12306.

- 1 <http://doi.wiley.com/10.1111/gcb.12306> (Accessed October 2, 2018).
- 2 Magnan, A. K., and Coauthors, 2016: Addressing the risk of maladaptation to climate change. *Wiley*
3 *Interdiscip. Rev. Clim. Chang.*, **7**, 646–665, doi:10.1002/wcc.409.
- 4 Mander, S., K. Anderson, A. Larkin, C. Gough, and N. Vaughan, 2017: The Role of Bio-energy with
5 Carbon Capture and Storage in Meeting the Climate Mitigation Challenge: A Whole System
6 Perspective. *Energy Procedia*, **114**, 6036–6043, doi:10.1016/J.EGYPRO.2017.03.1739.
7 <https://www.sciencedirect.com/science/article/pii/S1876610217319410> (Accessed March 16,
8 2018).
- 9 Marino, E., and H. Lazrus, 2015: Migration or Forced Displacement?: The Complex Choices of
10 Climate Change and Disaster Migrants in Shishmaref, Alaska and Nanumea, Tuvalu. *Hum.*
11 *Organ.*, **74**, 341–350, doi:10.17730/0018-7259-74.4.341.
- 12 Marjani, A., and M. Jamali, 2014: Role of exchange flow in salt water balance of Urmia Lake. *Dyn.*
13 *Atmos. Ocean.*, **65**, 1–16, doi:10.1016/j.dynatmoce.2013.10.001.
- 14 Martin-Ortega, J., T. E. H. Allott, K. Glenk, and M. Schaafsma, 2014: Valuing water quality
15 improvements from peatland restoration: Evidence and challenges. *Ecosyst. Serv.*,
16 doi:10.1016/j.ecoser.2014.06.00.
- 17 Masera, O. R., R. Bailis, R. Drigo, A. Ghilardi, and I. Ruiz-Mercado, 2015: Environmental Burden of
18 Traditional Bioenergy Use. *Annu. Rev. Environ. Resour.*, **40**, 121–150, doi:10.1146/annurev-
19 environ-102014-021318. [http://www.annualreviews.org/doi/10.1146/annurev-environ-102014-
20 021318](http://www.annualreviews.org/doi/10.1146/annurev-environ-102014-021318) (Accessed November 2, 2018).
- 21 Matata, P. Z., L. W. Masolwa, S. Ruvuga, and F. M. Bagarama, 2013: Dissemination pathways for
22 scaling-up agroforestry technologies in western Tanzania. *J. Agric. Ext. Rural Dev.*, **5**, 31–36.
- 23 Mateos, E., J. M. Edeso, and L. Ormaetxea, 2017: Soil Erosion and Forests Biomass as Energy
24 Resource in the Basin of the Oka River in Biscay, Northern Spain. *Forests*, **8**, 258,
25 doi:10.3390/f8070258.
- 26 Matthews, R., G. Hogan, and E. Mackie, 2018: Carbon impacts of biomass consumed in the EU. *Res.*
27 *Agency For. Comm.*, 61. [https://europeanclimate.org/wp-content/uploads/2018/05/CIB-
28 Summary-report-for-ECF-v10.5-May-20181.pdf](https://europeanclimate.org/wp-content/uploads/2018/05/CIB-Summary-report-for-ECF-v10.5-May-20181.pdf).
- 29 Mbow, C., P. Smith, D. Skole, L. Duguma, and M. Bustamante, 2014: Achieving mitigation and
30 adaptation to climate change through sustainable agroforestry practices in Africa. *Curr. Opin.*
31 *Environ. Sustain.*, **6**, 8–14, doi:10.1016/j.cosust.2013.09.002.
32 <https://www.sciencedirect.com/science/article/pii/S1877343513001255> (Accessed May 21,
33 2018).
- 34 McDowell, N. G., D. J. Beerling, D. D. Breshears, R. A. Fisher, K. F. Raffa, and M. Stitt, 2011: The
35 interdependence of mechanisms underlying climate-driven vegetation mortality. *Trends Ecol.*
36 *Evol.*, **26**, 523–532, doi:10.1016/J.TREE.2011.06.003.
37 <https://www.sciencedirect.com/science/article/pii/S0169534711001698> (Accessed March 5,
38 2018).
- 39 McGrath, M. J., and Coauthors, 2015: Reconstructing European forest management from 1600 to
40 2010. *Biogeosciences*, **12**, 4291–4316, doi:10.5194/bg-12-4291-2015.
- 41 McInnes, K. L., T. A. Erwin, and J. M. Bathols, 2011: Global Climate Model projected changes in 10
42 m wind speed and direction due to anthropogenic climate change. *Atmos. Sci. Lett.*, **12**, 325–333,
43 doi:10.1002/asl.341. <http://doi.wiley.com/10.1002/asl.341> (Accessed March 6, 2018).
- 44 McIntire, J., D. (Daniel) Bourzat, and P. L. Pingali, 1992: *Crop-livestock interaction in Sub-Saharan*
45 *Africa*. World Bank, 246 pp. <https://vtechworks.lib.vt.edu/handle/10919/65734> (Accessed April
46 2, 2019).
- 47 McLauchlan, K., 2006: The Nature and Longevity of Agricultural Impacts on Soil Carbon and
48 Nutrients: A Review. *Ecosystems*, **9**, 1364–1382, doi:10.1007/s10021-005-0135-1.

- 1 <http://link.springer.com/10.1007/s10021-005-0135-1> (Accessed May 21, 2018).
- 2 McLeman, R., 2017: *Migration and land degradation: Recent experience and future trends. Working*
3 *paper for the Global Land Outlook, 1st edition.* UNCCD, Bonn, 45 pp.
- 4 ———, and B. Smit, 2006: Migration as an Adaptation to Climate Change. *Clim. Change*, **76**, 31–53,
5 doi:10.1007/s10584-005-9000-7. <http://link.springer.com/10.1007/s10584-005-9000-7>
6 (Accessed October 30, 2018).
- 7 McLeman, R. A., J. Dupre, L. Berrang Ford, J. Ford, K. Gajewski, and G. Marchildon, 2014: What
8 we learned from the Dust Bowl: lessons in science, policy, and adaptation. *Popul. Environ.*, **35**,
9 417–440, doi:10.1007/s11111-013-0190-z. <http://link.springer.com/10.1007/s11111-013-0190-z>
10 (Accessed October 30, 2018).
- 11 McNicol, G., and W. L. Silver, 2014: Separate effects of flooding and anaerobiosis on soil greenhouse
12 gas emissions and redox sensitive biogeochemistry. *J. Geophys. Res. Biogeosciences*, **119**, 557–
13 566, doi:10.1002/2013JG002433.
14 <https://agupubs.onlinelibrary.wiley.com/doi/pdf/10.1002/2013JG002433> (Accessed May 8,
15 2018).
- 16 McNicol, I. M., C. M. Ryan, and E. T. A. Mitchard, 2018: Carbon losses from deforestation and
17 widespread degradation offset by extensive growth in African woodlands. *Nat. Commun.*, **9**,
18 3045, doi:10.1038/s41467-018-05386-z. <https://doi.org/10.1038/s41467-018-05386-z>.
- 19 McVicar, T. R., and M. L. Roderick, 2010: Winds of change. *Nat. Geosci.*, **3**, 747–748,
20 doi:10.1038/ngeo1002. <http://www.nature.com/articles/ngeo1002> (Accessed March 8, 2019).
- 21 Meijer, S. S., D. Catacutan, O. C. Ajayi, G. W. Sileshi, and M. Nieuwenhuis, 2015: The role of
22 knowledge, attitudes and perceptions in the uptake of agricultural and agroforestry innovations
23 among smallholder farmers in sub-Saharan Africa. *Int. J. Agric. Sustain.*, **13**, 40–54,
24 doi:10.1080/14735903.2014.912493.
25 <https://www.tandfonline.com/doi/full/10.1080/14735903.2014.912493> (Accessed March 25,
26 2019).
- 27 Meisner, A., S. Jacquiod, B. L. Snoek, F. C. ten Hooven, and W. H. van der Putten, 2018: Drought
28 Legacy Effects on the Composition of Soil Fungal and Prokaryote Communities. *Front.*
29 *Microbiol.*, **9**, 294, doi:10.3389/fmicb.2018.00294.
30 <http://journal.frontiersin.org/article/10.3389/fmicb.2018.00294/full> (Accessed March 6, 2019).
- 31 Mekuriaw, A., A. Heinimann, G. Zeleke, and H. Hurni, 2018: Factors influencing the adoption of
32 physical soil and water conservation practices in the Ethiopian highlands. *Int. Soil Water*
33 *Conserv. Res.*, **6**, 23–30, doi:10.1016/J.ISWCR.2017.12.006.
34 <https://www.sciencedirect.com/science/article/pii/S2095633917302058> (Accessed March 25,
35 2019).
- 36 Mendum, R., and M. Njenga, 2018: *Recovering bioenergy in Sub-Saharan Africa: gender dimensions,*
37 *lessons and Challenges.* International Water Management Institute (IWMI), Colombo, Sri Lanka,
38 1–83 pp.
39 [https://books.google.se/books?hl=sv&lr=&id=Nm18DwAAQBAJ&oi=fnd&pg=PA1&dq=biom
40 ass+cooking+women+youth&ots=IA8ciTkXzH&sig=sbEH1LUT_FXDseWOt-
41 fC98tL8fY&redir_esc=y#v=onepage&q=biomass cooking women youth&f=false](https://books.google.se/books?hl=sv&lr=&id=Nm18DwAAQBAJ&oi=fnd&pg=PA1&dq=biomass+cooking+women+youth&ots=IA8ciTkXzH&sig=sbEH1LUT_FXDseWOt-fC98tL8fY&redir_esc=y#v=onepage&q=biomass%20cooking%20women%20youth&f=false) (Accessed
42 April 10, 2019).
- 43 Meng, W., M. He, B. Hu, X. Mo, H. Li, B. Liu, and Z. Wang, 2017: Status of wetlands in China: A
44 review of extent, degradation, issues and recommendations for improvement. *Ocean Coast.*
45 *Manag.*, **146**, 50–59, doi:10.1016/J.OCECOAMAN.2017.06.003.
46 [https://www.sciencedirect.com/science/article/pii/S0964569117305550?_rdoc=1&_fmt=high&
47 _origin=gateway&_docanchor=&md5=b8429449ccfc9c30159a5f9aeaa92ffb&dgcid=raven_sd_re
48 commender_email#bib50](https://www.sciencedirect.com/science/article/pii/S0964569117305550?_rdoc=1&_fmt=high&_origin=gateway&_docanchor=&md5=b8429449ccfc9c30159a5f9aeaa92ffb&dgcid=raven_sd_recommender_email#bib50) (Accessed July 18, 2018).
- 49 Mengel, M., A. Levermann, K. Frieler, A. Robinson, B. Marzeion, and R. Winkelmann, 2016: Future

- 1 sea level rise constrained by observations and long-term commitment. *Proc. Natl. Acad. Sci.*,
2 **113**, 2597–2602, doi:10.1073/PNAS.1500515113.
3 <https://www.pnas.org/content/113/10/2597.short> (Accessed March 14, 2019).
- 4 Mentaschi, L., M. I. Vousdoukas, J.-F. Pekel, E. Voukouvalas, and L. Feyen, 2018: Global long-term
5 observations of coastal erosion and accretion. *Sci. Rep.*, **8**, 12876, doi:10.1038/s41598-018-
6 30904-w. <http://www.nature.com/articles/s41598-018-30904-w> (Accessed October 1, 2018).
- 7 Mercer, D. E., 2004: Adoption of agroforestry innovations in the tropics: A review. *Agrofor. Syst.*,
8 **61–62**, 311–328, doi:10.1023/B:AGFO.0000029007.85754.70.
9 <http://link.springer.com/10.1023/B:AGFO.0000029007.85754.70> (Accessed October 28, 2018).
- 10 Meyfroidt, P., and E. F. Lambin, 2011: *Global Forest Transition: Prospects for an End to*
11 *Deforestation*. 343–371 pp. [http://www.annualreviews.org/doi/10.1146/annurev-environ-
12 *090710-143732*.](http://www.annualreviews.org/doi/10.1146/annurev-environ-090710-143732)
- 13 Micklin, P., 2010: The past, present, and future Aral Sea. *Lakes Reserv. Res. Manag.*, **15**, 193–213,
14 doi:10.1111/j.1440-1770.2010.00437.x. <http://doi.wiley.com/10.1111/j.1440-1770.2010.00437.x>
15 (Accessed December 27, 2017).
- 16 Middleton, N. J., and D. S. Thomas, 1997: *World Atlas of Desertification*.
17 <http://www.ncbi.nlm.nih.gov/pubmed/12345678> (Accessed May 25, 2018).
- 18 Midgley, G. F., and W. J. Bond, 2015: Future of African terrestrial biodiversity and ecosystems under
19 anthropogenic climate change. *Nat. Clim. Chang.*, **5**, 823–829, doi:10.1038/nclimate2753.
20 <http://www.nature.com/articles/nclimate2753> (Accessed November 1, 2018).
- 21 Miettinen, J., C. Shi, and S. C. Liew, 2016: Land cover distribution in the peatlands of Peninsular
22 Malaysia, Sumatra and Borneo in 2015 with changes since 1990. *Glob. Ecol. Conserv.*, **6**, 67–
23 78, doi:10.1016/j.gecco.2016.02.004.
- 24 Miles, L., and Coauthors, 2015: *Mitigation potential from forest-related activities and incentives for*
25 *enhanced action in developing countries*. United Nations Environment Programme, Nairobi,
26 Kenya, <http://www.forskningssdatabasen.dk/en/catalog/2290491813> (Accessed November 1,
27 2018).
- 28 Millennium Assessment, 2005: *Ecosystems and Human Well-being, Synthesis*. Island Press,
29 Washington DC, USA, 155 pp.
- 30 Minasny, B., and Coauthors, 2017: Soil carbon 4 per mille. *Geoderma*, **292**, 59–86,
31 doi:10.1016/J.GEODERMA.2017.01.002.
32 <https://www.sciencedirect.com/science/article/pii/S0016706117300095> (Accessed October 3,
33 2018).
- 34 Minderhoud, P. S. J., G. Erkens, V. H. Pham, V. T. Bui, L. Erban, H. Kooi, and E. Stouthamer, 2017:
35 Impacts of 25 years of groundwater extraction on subsidence in the Mekong delta, Vietnam.
36 *Environ. Res. Lett.*, **12**, 064006, doi:10.1088/1748-9326/aa7146. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=6/a=064006?key=crossref.687693ccba9bba2f8d848aa7018a5625)
37 [9326/12/i=6/a=064006?key=crossref.687693ccba9bba2f8d848aa7018a5625](http://stacks.iop.org/1748-9326/12/i=6/a=064006?key=crossref.687693ccba9bba2f8d848aa7018a5625) (Accessed October
38 27, 2018).
- 39 Minkinen, K., P. Ojanen, T. Penttilä, M. Aurela, T. Laurila, J.-P. Tuovinen, and A. Lohila, 2018:
40 Persistent carbon sink at a boreal drained bog forest. *Biogeosciences*, **15**, 3603–3624,
41 doi:10.5194/bg-15-3603-2018.
- 42 Minx, J. C., and Coauthors, 2018: Negative emissions - Part 1: Research landscape and synthesis.
43 *Environ. Res. Lett.*, **13**, doi:10.1088/1748-9326/aabf9b.
- 44 Mirzabaev, A., E. Nkonya, and J. von Braun, 2015: Economics of sustainable land management.
45 *Curr. Opin. Environ. Sustain.*, **15**, 9–19, doi:10.1016/J.COSUST.2015.07.004.
46 <https://www.sciencedirect.com/science/article/pii/S1877343515000688> (Accessed April 18,
47 2019).
- 48 —, —, J. Goedecke, T. Johnson, and W. Anderson, 2016: Global Drivers of Land Degradation

- 1 and Improvement. *Economics of Land Degradation and Improvement – A Global Assessment for*
2 *Sustainable Development*, Springer International Publishing, Cham, 167–195
3 http://link.springer.com/10.1007/978-3-319-19168-3_7 (Accessed January 12, 2018).
- 4 Mitchard, E. T. A., 2018: The tropical forest carbon cycle and climate change. *Nature*, **559**, 527–534,
5 doi:10.1038/s41586-018-0300-2.
- 6 Mittler, R., 2006: Abiotic stress, the field environment and stress combination. *Trends Plant Sci.*, **11**,
7 15–19, doi:10.1016/J.TPLANTS.2005.11.002.
8 <https://www.sciencedirect.com/science/article/pii/S1360138505002918> (Accessed May 23,
9 2018).
- 10 Mokria, M., A. Gebrekirstos, E. Aynekulu, and A. Bräuning, 2015: Tree dieback affects climate
11 change mitigation potential of a dry afro-montane forest in northern Ethiopia. *For. Ecol.*
12 *Manage.*, **344**, 73–83, doi:10.1016/j.foreco.2015.02.008.
13 <https://linkinghub.elsevier.com/retrieve/pii/S0378112715000596> (Accessed October 27, 2018).
- 14 Mondal, A., D. Khare, and S. Kundu, 2016: Change in rainfall erosivity in the past and future due to
15 climate change in the central part of India. *Int. Soil Water Conserv. Res.*, **4**, 186–194,
16 doi:10.1016/J.ISWCR.2016.08.004.
17 <http://www.sciencedirect.com/science/article/pii/S2095633916300053> (Accessed December 5,
18 2017).
- 19 Montanarella, L., R. Scholes, and A. Brainich, 2018: *The IPBES assessment report on land*
20 *degradation and restoration*. Bonn, Germany,.
- 21 Montgomery, D. R., 2007a: Soil erosion and agricultural sustainability. **104**, 13268–13272.
22 <http://www.pnas.org/content/pnas/104/33/13268.full.pdf> (Accessed April 2, 2018).
- 23 Montgomery, D. R., 2007b: *Dirt: the Erosion of Civilizations*. University of California Press, 285 pp.
- 24 Moore, P. D., 2002: *The future of cool temperate bogs*. *Environmental Conservation*. 29, 3-20.
25 *National Climate Commission.2009. Belgium’s Fifth National Communication under the United*
26 *Nations Framework Convention on Climate Change. Federal Public Service Health, Food Chain*
27 *Saft*.
- 28 Moore, S., and Coauthors, 2013: Deep instability of deforested tropical peatlands revealed by fluvial
29 organic carbon fluxes. *Nature*, **493**, 660–663, doi:10.1038/nature11818.
- 30 Morgan, R. P. C., 2005a: *Soil Erosion and Conservation*. 3rd ed. Blackwell Science Ltd, Malden,
31 USA,.
- 32 Morgan, R. P. C. (Royston P. C., 2005b: *Soil erosion and conservation*. 2nd ed. Blackwell Pub,
33 Harlow, Essex, UK, 198 pp.
- 34 Mori, N., T. Yasuda, H. Mase, T. Tom, and Y. Oku, 2010: Projection of Extreme Wave Climate
35 Change under Global Warming. *Hydrol. Res. Lett.*, **4**, 15–19, doi:10.3178/hr1.4.15.
36 https://www.jstage.jst.go.jp/article/hr1/4/0/4_0_15/_article (Accessed March 5, 2018).
- 37 Morrissey, J. W., 2013: Understanding the relationship between environmental change and migration:
38 The development of an effects framework based on the case of northern Ethiopia. *Glob.*
39 *Environ. Chang.*, **23**, 1501–1510, doi:10.1016/J.GLOENVCHA.2013.07.021.
40 <https://www.sciencedirect.com/science/article/pii/S0959378013001271> (Accessed May 15,
41 2018).
- 42 Morse, J. L., and E. S. Bernhardt, 2013: Using ¹⁵N tracers to estimate N₂O and N₂ emissions from
43 nitrification and denitrification in coastal plain wetlands under contrasting land-uses. *Soil Biol.*
44 *Biochem.*, doi:10.1016/j.soilbio.2012.07.025.
- 45 Morton, J. F., 2007: The impact of climate change on smallholder and subsistence agriculture. *Proc.*
46 *Natl. Acad. Sci. U. S. A.*, **104**, 19680–19685, doi:10.1073/pnas.0701855104.
47 <http://www.ncbi.nlm.nih.gov/pubmed/18077400> (Accessed January 9, 2018).

- 1 Mudombi, S., and Coauthors, 2018: Multi-dimensional poverty effects around operational biofuel
2 projects in Malawi, Mozambique and Swaziland. *Biomass and Bioenergy*, **114**, 41–54,
3 doi:<https://doi.org/10.1016/j.biombioe.2016.09.003>.
4 <http://www.sciencedirect.com/science/article/pii/S0961953416302938>.
- 5 Muir, D., J. A. G. Cooper, and G. Pétursdóttir, 2014: Challenges and opportunities in climate change
6 adaptation for communities in Europe’s northern periphery. *Ocean Coast. Manag.*, **94**, 1–8,
7 doi:10.1016/j.ocecoaman.2014.03.017.
- 8 Mullan, D., 2013: Soil erosion under the impacts of future climate change: Assessing the statistical
9 significance of future changes and the potential on-site and off-site problems. *CATENA*, **109**,
10 234–246, doi:10.1016/J.CATENA.2013.03.007.
11 <https://www.sciencedirect.com/science/article/pii/S0341816213000696> (Accessed March 15,
12 2018).
- 13 —, D. Favis-Mortlock, and R. Fealy, 2012: Addressing key limitations associated with modelling
14 soil erosion under the impacts of future climate change. *Agric. For. Meteorol.*, **156**, 18–30,
15 doi:10.1016/j.agrformet.2011.12.004.
16 <http://www.sciencedirect.com/science/article/pii/S0168192311003418> (Accessed November 15,
17 2017).
- 18 Munang, R., I. Thiaw, K. Alverson, and Z. Han, 2013: The role of ecosystem services in climate
19 change adaptation and disaster risk reduction. *Curr. Opin. Environ. Sustain.*, **5**, 47–52,
20 doi:10.1016/J.COSUST.2013.02.002.
21 <https://www.sciencedirect.com/science/article/pii/S1877343513000080> (Accessed May 17,
22 2018).
- 23 Munson, S. M., J. Belnap, and G. S. Okin, 2011: Responses of wind erosion to climate-induced
24 vegetation changes on the Colorado Plateau. *Proc. Natl. Acad. Sci.*, **108**, 3854–3859,
25 doi:10.1073/pnas.1014947108. <http://www.pnas.org/cgi/doi/10.1073/pnas.1014947108>
26 (Accessed November 14, 2017).
- 27 Murphy, J. M., D. M. H. Sexton, D. N. Barnett, G. S. Jones, M. J. Webb, M. Collins, and D. A.
28 Stainforth, 2004: Quantification of modelling uncertainties in a large ensemble of climate
29 change simulations. *Nature*, **430**, 768–772, doi:10.1038/nature02771.
30 <http://www.nature.com/doi/10.1038/nature02771> (Accessed May 21, 2018).
- 31 Murthy, K., and S. Bagchi, 2018: Spatial patterns of long-term vegetation greening and browning are
32 consistent across multiple scales: Implications for monitoring land degradation. *L. Degrad. Dev.*,
33 **29**, 2485–2495, doi:10.1002/ldr.3019. <http://doi.wiley.com/10.1002/ldr.3019> (Accessed October
34 10, 2018).
- 35 Murty, D., M. U. F. Kirschbaum, R. E. Mcmurtrie, and H. Mcgilvray, 2002: Does conversion of forest
36 to agricultural land change soil carbon and nitrogen? a review of the literature. *Glob. Chang.*
37 *Biol.*, **8**, 105–123, doi:10.1046/j.1354-1013.2001.00459.x.
- 38 Mutoko, M. C., C. A. Shisanya, and L. Hein, 2014: Fostering technological transition to sustainable
39 land management through stakeholder collaboration in the western highlands of Kenya. *Land*
40 *use policy*, **41**, 110–120, doi:10.1016/J.LANDUSEPOL.2014.05.005.
41 <https://www.sciencedirect.com/science/article/pii/S0264837714001112?via%3Dihub> (Accessed
42 November 2, 2018).
- 43 Mycoo, M., and A. Chadwick, 2012: Adaptation to climate change: the coastal zone of Barbados.
44 *Proc. Inst. Civ. Eng. - Marit. Eng.*, **165**, 159–168, doi:10.1680/maen.2011.19.
- 45 Naamwintome, B. A., and E. Bagson, 2013: Youth in agriculture: Prospects and challenges in the
46 Sissala area of Ghana. *Net J. Agric. Sci.*, **1**, 60–68.
47 http://www.netjournals.org/z_NJAS_13_023.html (Accessed April 6, 2019).
- 48 Nabuurs, G. J., and Coauthors, 2007: In Climate Change 2007: Mitigation. Contribution of Working
49 Group III to the Fourth Assessment Report of the Intergovernmental Panel on Climate Change.

- 1 *Forestry*, 542–584.
- 2 Nabuurs, G. J., P. Delacote, D. Ellison, M. Hanewinkel, L. Hetemäki, M. Lindner, and M. Ollikainen,
3 2017: By 2050 the mitigation effects of EU forests could nearly double through climate smart
4 forestry. *Forests*, **8**, 1–14, doi:10.3390/f8120484.
- 5 Nachtergaele, F. O., M. Petri, R. Biancalani, G. V. Lynden, H. V. Velthuizen, and M. Bloise, 2011:
6 *Nachtergaele, F. O., et al. Global Land Degradation Information System (GLADIS) Version 1.0.*
7 *An Information database for Land Degradation Assessment at Global Level. LADA Technical*
8 *Report 17.* Rome, Italy, 1-110 pp.
- 9 Nalau, J., J. Handmer, J. Nalau, and J. Handmer, 2018: Improving Development Outcomes and
10 Reducing Disaster Risk through Planned Community Relocation. *Sustainability*, **10**, 3545,
11 doi:10.3390/su10103545. <http://www.mdpi.com/2071-1050/10/10/3545> (Accessed March 30,
12 2019).
- 13 Nampak, H., B. Pradhan, H. Mojaddadi Rizeei, and H.-J. Park, 2018: Assessment of land cover and
14 land use change impact on soil loss in a tropical catchment by using multitemporal SPOT-5
15 satellite images and Revised Universal Soil Loss Equation model. *L. Degrad. Dev.*,
16 doi:10.1002/ldr.3112. <http://doi.wiley.com/10.1002/ldr.3112> (Accessed October 1, 2018).
- 17 Nasi, R., F. E. Putz, P. Pacheco, S. Wunder, and S. Anta, 2011: Sustainable forest management and
18 carbon in tropical Latin America: The case for REDD+. *Forests*, **2**, 200–217,
19 doi:10.3390/f2010200.
- 20 Nath, T. K., M. Jashimuddin, M. Kamrul Hasan, M. Shahjahan, and J. Pretty, 2016: The sustainable
21 intensification of agroforestry in shifting cultivation areas of Bangladesh. *Agrofor. Syst.*, **90**,
22 405–416, doi:10.1007/s10457-015-9863-1. <http://link.springer.com/10.1007/s10457-015-9863-1>
23 (Accessed May 21, 2018).
- 24 National Research Centre for Agroforestry, J., and Coauthors, 1999: *Indian journal of agroforestry.*
25 Indian Society for Agroforestry, Research Centre for Agroforestry, 1-10 pp.
26 <http://www.indianjournals.com/ijor.aspx?target=ijor:ijaf&volume=20&issue=1&article=001>
27 (Accessed October 22, 2018).
- 28 Ndegwa, G. M., U. Nehren, F. Grüniger, M. Iiyama, and D. Anhof, 2016: Charcoal production
29 through selective logging leads to degradation of dry woodlands: a case study from Mutomo
30 District, Kenya. *J. Arid Land*, **8**, 618–631, doi:10.1007/s40333-016-0124-6.
31 <https://doi.org/10.1007/s40333-016-0124-6>.
- 32 Nearing, M. A., F. F. Pruski, and M. R. O’Neal, 2004a: Expected climate change impacts on soil
33 erosion rates: A review. *J. Soil Water Conserv.*, **59**, 43–50.
34 <http://www.jswnonline.org/content/59/1/43.short> (Accessed December 5, 2017).
- 35 Nearing, M. A., F. F. Pruski, and M. R. O’Neal, 2004b: Expected climate change impacts on
36 soil erosion rates: A review. **59**, 43–50. <https://www.scopus.com/record/display.uri?eid=2-s2.0-0742319939&origin=resultslist&sort=cp-f&src=s&st1=%22climate+change%22+and+soil+erosion+and+%28water+or+hydric%29&nlo=&nlr=&nls=&sid=f321ece3d740bda383b499527e324fdb&sot=b&sdt=b&sl=70&s=TITLE-ABS-KEY%28> (Accessed February 14, 2018).
- 41 Nearing, M. A., and Coauthors, 2005: Modeling response of soil erosion and runoff to changes in
42 precipitation and cover. *CATENA*, **61**, 131–154, doi:10.1016/j.catena.2005.03.007.
43 <http://www.sciencedirect.com/science/article/pii/S0341816205000512> (Accessed November 14,
44 2017).
- 45 Nemet, G. F., and Coauthors, 2018: Negative emissions - Part 3: Innovation and upscaling. *Environ.*
46 *Res. Lett.*, **13**, doi:10.1088/1748-9326/aabff4.
- 47 Nesshöver, C., and Coauthors, 2017: The science, policy and practice of nature-based solutions: An
48 interdisciplinary perspective. *Sci. Total Environ.*, **579**, 1215–1227,
49 doi:10.1016/J.SCITOTENV.2016.11.106.

- 1 <https://www.sciencedirect.com/science/article/pii/S0048969716325578> (Accessed April 3,
2 2019).
- 3 Netting, R. M., 1993: *Smallholders, householders : farm families and the ecology of intensive,*
4 *sustainable agriculture*. Stanford University Press, 389 pp.
- 5 Neumann, B., A. T. Vafeidis, J. Zimmermann, and R. J. Nicholls, 2015: Future Coastal Population
6 Growth and Exposure to Sea-Level Rise and Coastal Flooding - A Global Assessment. *PLoS*
7 *One*, **10**, e0118571, doi:10.1371/journal.pone.0118571.
8 <https://dx.plos.org/10.1371/journal.pone.0118571> (Accessed October 27, 2018).
- 9 Neupane, R. P., and S. Kumar, 2015: Estimating the effects of potential climate and land use changes
10 on hydrologic processes of a large agriculture dominated watershed. *J. Hydrol.*, **529**, 418–429,
11 doi:10.1016/J.JHYDROL.2015.07.050.
12 <https://www.sciencedirect.com/science/article/pii/S0022169415005569> (Accessed March 14,
13 2018).
- 14 Newton, P., J. A. Oldekop, G. Brodnig, B. K. Karna, and A. Agrawal, 2016: Carbon, biodiversity, and
15 livelihoods in forest commons: synergies, trade-offs, and implications for REDD+. *Environ. Res.*
16 *Let.*, **11**, 044017, doi:10.1088/1748-9326/11/4/044017. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=4/a=044017?key=crossref.e432d29d21198b79ca24de2b3805cbde)
17 [9326/11/i=4/a=044017?key=crossref.e432d29d21198b79ca24de2b3805cbde](http://stacks.iop.org/1748-9326/11/i=4/a=044017?key=crossref.e432d29d21198b79ca24de2b3805cbde) (Accessed May 27,
18 2018).
- 19 Nguyen, T. T. N., C. Y. Xu, I. Tahmasbian, R. Che, Z. Xu, X. Zhou, H. M. Wallace, and S. H. Bai,
20 2017: Effects of biochar on soil available inorganic nitrogen: A review and meta-analysis.
21 *Geoderma*, doi:10.1016/j.geoderma.2016.11.004.
- 22 Nicholls, R. J., N. Marinova, J. A. Lowe, S. Brown, P. Vellinga, D. de Gusmão, J. Hinkel, and R. S. J.
23 Tol, 2011: Sea-level rise and its possible impacts given a “beyond 4°C world” in the twenty-first
24 century. *Philos. Trans. A. Math. Phys. Eng. Sci.*, **369**, 161–181, doi:10.1098/rsta.2010.0291.
25 <http://www.ncbi.nlm.nih.gov/pubmed/21115518> (Accessed May 31, 2018).
- 26 Nicholls, R. J., C. Woodroffe, and V. Burkett, 2016: Coastline Degradation as an Indicator of Global
27 Change. *Clim. Chang.*, 309–324, doi:10.1016/B978-0-444-63524-2.00020-8.
28 <https://www.sciencedirect.com/science/article/pii/B9780444635242000208> (Accessed March
29 16, 2019).
- 30 —, and Coauthors, 2018: Stabilization of global temperature at 1.5°C and 2.0°C: implications for
31 coastal areas. *Philos. Trans. R. Soc. A Math. Eng. Sci.*, **376**, 20160448,
32 doi:10.1098/rsta.2016.0448.
33 <http://rsta.royalsocietypublishing.org/lookup/doi/10.1098/rsta.2016.0448> (Accessed March 14,
34 2019).
- 35 Niedertscheider, M., T. Kastner, T. Fetzl, H. Haberl, C. Kroisleitner, C. Plutzer, and K.-H. Erb,
36 2016: Mapping and analysing cropland use intensity from a NPP perspective. *Environ. Res.*
37 *Let.*, **11**, 014008, doi:10.1088/1748-9326/11/1/014008. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=1/a=014008?key=crossref.d5afd2fdaa12ac43b37c6c621dd3291a)
38 [9326/11/i=1/a=014008?key=crossref.d5afd2fdaa12ac43b37c6c621dd3291a](http://stacks.iop.org/1748-9326/11/i=1/a=014008?key=crossref.d5afd2fdaa12ac43b37c6c621dd3291a) (Accessed May 24,
39 2018).
- 40 Nielsen, D. L., and M. A. Brock, 2009: Modified water regime and salinity as a consequence of
41 climate change: prospects for wetlands of Southern Australia. *Clim. Change*, **95**, 523–533,
42 doi:10.1007/s10584-009-9564-8. <http://link.springer.com/10.1007/s10584-009-9564-8>
43 (Accessed March 18, 2019).
- 44 Ntcheu Tchidje, S., D. J. Sonwa, B.-A. Nkongmeneck, L. Cerbonney, and R. Sufo Kankeu, 2016:
45 Preliminary estimation of carbon stock in a logging concession with a forest management plan in
46 East Cameroon. *J. Sustain. For.*, **35**, 355–368, doi:10.1080/10549811.2016.1190757.
47 <http://www.tandfonline.com/doi/full/10.1080/10549811.2016.1190757> (Accessed October 31,
48 2018).
- 49 Njenga, M., J. K. Gitau, M. Iiyama, R. Jamnadassa, Y. Mahmoud, and N. Karanja, 2019: innovative

- 1 biomass cooking approaches for sub-Saharan Africa. *African J. Food, Agric. Nutr. Dev.*, **19**,
2 14066–14087. <https://www.ajfand.net/Volume19/No1/BLFB1031.pdf>.
- 3 Nkonya, E., M. Winslow, M. S. Reed, M. Mortimore, and A. Mirzabaev, 2011: Monitoring and
4 assessing the influence of social, economic and policy factors on sustainable land management
5 in drylands. *L. Degrad. Dev.*, **22**, 240–247, doi:10.1002/ldr.1048.
6 <http://doi.wiley.com/10.1002/ldr.1048> (Accessed November 1, 2018).
- 7 Nkonya, E., A. Mirzabaev, and J. von Braun, 2016: *Economics of Land Degradation and*
8 *Improvement - A Global Assessment for Sustainable Development*. E. Nkonya, A. Mirzabaev,
9 and J. Von Braun, Eds. Springer, Heidelberg, New York, Dordrecht, London, 695 pp.
10 <http://www.awdflibrary.org/handle/123456789/289> (Accessed March 15, 2019).
- 11 Noordin, Q., A. Niang, B. Jama, and M. Nyasimi, 2001: Scaling up adoption and impact of
12 agroforestry technologies: Experiences from western Kenya. *Dev. Pract.*, **11**, 509–523,
13 doi:10.1080/09614520120066783.
14 <http://www.tandfonline.com/doi/abs/10.1080/09614520120066783> (Accessed October 22,
15 2018).
- 16 Noormets, A., D. Epron, J. C. Domec, S. G. McNulty, T. Fox, G. Sun, and J. S. King, 2015: Effects of
17 forest management on productivity and carbon sequestration : A review and hypothesis q. *For.*
18 *Ecol. Manage.*, **355**, 124–140, doi:10.1016/j.foreco.2015.05.019.
19 <http://dx.doi.org/10.1016/j.foreco.2015.05.019>.
- 20 Norby, R. J., J. M. Warren, C. M. Iversen, B. E. Medlyn, and R. E. McMurtrie, 2010: CO2
21 enhancement of forest productivity constrained by limited nitrogen availability. *Proc. Natl.*
22 *Acad. Sci. USA*, **107**, 19368–19373, doi:10.1073/pnas.1006463107/
23 /DCSupplemental.www.pnas.org/cgi/doi/10.1073/pnas.1006463107.
- 24 Norton, J. B., L. J. Jungst, U. Norton, H. R. Olsen, K. W. Tate, and W. R. Horwath, 2011: Soil
25 Carbon and Nitrogen Storage in Upper Montane Riparian Meadows. *Ecosystems*, **14**, 1217–
26 1231, doi:10.1007/s10021-011-9477-z. <http://link.springer.com/10.1007/s10021-011-9477-z>
27 (Accessed February 14, 2018).
- 28 Nugent, K. A., I. B. Strachan, M. Strack, N. T. Roulet, and L. Rochefort, 2018: Multi-year net
29 ecosystem carbon balance of a restored peatland reveals a return to carbon sink. *Glob. Chang.*
30 *Biol.*, doi:10.1111/gcb.14449.
- 31 Nunes, J. P., and M. A. Nearing, 2011: Modelling Impacts of Climate Change: Case studies using the
32 new generation of erosion models. *Handbook of Erosion Modelling*, R.P.C. Morgan and M.A.
33 Nearing, Eds., Wiley, Chichester, West Sussex, UK, p. 400.
- 34 Nurse, L. A., R. F. McLean, J. Agard, L. Briguglio, V. Duvat-Magnan, N. Pelesikoti, E. L. Tompkins,
35 and A. P. Webb, 2014: Small islands. *Climate Change 2014: Impacts, Adaptation, and*
36 *Vulnerability*, V.R. Barros et al., Eds., Cambridge University Press, Cambridge; New York,
37 1613–1654.
- 38 Nyland, R. D., 1992: Exploitation and Greed in Eastern Hardwood Forests. *J. For.*, **90**, 33–37,
39 doi:10.1093/jof/90.1.33. <https://academic.oup.com/jof/article-abstract/90/1/33/4635445>
40 (Accessed May 31, 2018).
- 41 O'Brien, K., and Coauthors, 2004: Mapping vulnerability to multiple stressors: climate change and
42 globalization in India. *Glob. Environ. Chang.*, **14**, 303–313,
43 doi:10.1016/J.GLOENVCHA.2004.01.001.
44 <https://www.sciencedirect.com/science/article/pii/S095937800400010X> (Accessed May 14,
45 2018).
- 46 O'Connell, D., and Coauthors, 2016: Designing projects in a rapidly changing world: guidelines for
47 embedding resilience, adaptation and transformation into sustainable development projects.
48 *Glob. Environ. Facil. Washingt.*, <http://www.stagef.org/rapta-guidelines>.
- 49 O'Connor, D., T. Peng, J. Zhang, D. C. W. Tsang, D. S. Alessi, Z. Shen, N. S. Bolan, and D. Hou,

- 1 2018: Biochar application for the remediation of heavy metal polluted land: A review of in situ
2 field trials. *Sci. Total Environ.*, **619–620**, 815–826, doi:10.1016/j.scitotenv.2017.11.132.
3 <http://linkinghub.elsevier.com/retrieve/pii/S004896971733190X> (Accessed December 2, 2017).
- 4 O’Driscoll, C., J. Sheahan, F. Renou-Wilson, P. Croot, F. Pilla, B. Misstear, and L. Xiao, 2018:
5 National scale assessment of total trihalomethanes in Irish drinking water. *J. Environ. Manage.*,
6 doi:10.1016/j.jenvman.2018.01.070.
- 7 O’Gorman, P. A., 2015: Precipitation Extremes Under Climate Change. *Curr. Clim. Chang. Reports*,
8 **1**, 49–59, doi:10.1007/s40641-015-0009-3. <http://link.springer.com/10.1007/s40641-015-0009-3>
9 (Accessed March 7, 2019).
- 10 O’Neal, M. R., M. A. Nearing, R. C. Vining, J. Southworth, and R. A. Pfeifer, 2005: Climate change
11 impacts on soil erosion in Midwest United States with changes in crop management. *CATENA*,
12 **61**, 165–184, doi:10.1016/J.CATENA.2005.03.003.
13 <https://www.sciencedirect.com/science/article/pii/S0341816205000421> (Accessed May 23,
14 2018).
- 15 Odgaard, M. V., M. T. Knudsen, J. E. Hermansen, and T. Dalgaard, 2019: Targeted grassland
16 production – A Danish case study on multiple benefits from converting cereal to grasslands for
17 green biorefinery. *J. Clean. Prod.*, **223**, 917–927, doi:10.1016/J.JCLEPRO.2019.03.072.
18 <https://www.sciencedirect.com/science/article/pii/S0959652619307632> (Accessed April 11,
19 2019).
- 20 Oertel, C., J. Matschullat, K. Zurba, F. Zimmermann, and S. Erasmi, 2016: Greenhouse gas emissions
21 from soils—A review. *Chemie der Erde - Geochemistry*, **76**, 327–352,
22 doi:10.1016/J.CHEMER.2016.04.002.
23 <https://www.sciencedirect.com/science/article/pii/S0009281916300551#bib0460> (Accessed May
24 12, 2018).
- 25 Ogle, S. M., A. Swan, and K. Paustian, 2012: No-till management impacts on crop productivity,
26 carbon input and soil carbon sequestration. *Agric. Ecosyst. Environ.*, **149**, 37–49,
27 doi:10.1016/J.AGEE.2011.12.010.
28 <https://www.sciencedirect.com/science/article/pii/S0167880911004361> (Accessed March 31,
29 2019).
- 30 Ojanen, P., A. Lehtonen, J. Heikkinen, T. Penttilä, and K. Minkkinen, 2014: Soil CO₂ balance and its
31 uncertainty in forestry-drained peatlands in Finland. *For. Ecol. Manage.*, **325**, 60–73,
32 doi:10.1016/j.foreco.2014.03.049.
- 33 Ojeda, G., J. Patrício, S. Mattana, and A. J. F. N. Sobral, 2016: Effects of biochar addition to estuarine
34 sediments. *J. Soils Sediments*, **16**, 2482–2491, doi:10.1007/s11368-016-1493-3.
35 <http://link.springer.com/10.1007/s11368-016-1493-3> (Accessed April 2, 2019).
- 36 Ojima, D. S., K. A. Galvin, and B. L. Turner, 1994: The Global Impact of Land-Use Change.
37 *Bioscience*, **44**, 300–304, doi:10.2307/1312379. [https://academic.oup.com/bioscience/article-](https://academic.oup.com/bioscience/article-lookup/doi/10.2307/1312379)
38 [lookup/doi/10.2307/1312379](https://academic.oup.com/bioscience/article-lookup/doi/10.2307/1312379) (Accessed October 3, 2018).
- 39 Oktarita, S., K. Hergoualc’h, S. Anwar, and L. V Verchot, 2017: Environmental Research Letters
40 Substantial N₂O emissions from peat decomposition and N fertilization in an oil palm
41 plantation exacerbated by hotspots Substantial N₂O emissions from peat decomposition and N
42 fertilization in an oil palm plantation exac. *Environ. Res. Lett.*, **12**.
- 43 Oldeman, L. R., and G. W. J. van Lynden, 1998: Revisiting the Glasod Methodology. *Methods for*
44 *Assessment of Soil Degradation*, R. Lal, W.H. Blum, C. Valentine, and B.A. Stewart, Eds., CRC
45 Press, Boca Raton, London, New York, Washington D.C., 423–440.
- 46 Olguin, M., and Coauthors, 2018: Applying a systems approach to assess carbon emission reductions
47 from climate change mitigation in Mexico’s forest sector. *Environ. Res. Lett.*, **13**,
48 doi:10.1088/1748-9326/aaaa03.
- 49 de Oliveira, G., N. A. Brunsell, C. E. Sutherlin, T. E. Crews, and L. R. DeHaan, 2018: Energy, water

- 1 and carbon exchange over a perennial Kernza wheatgrass crop. *Agric. For. Meteorol.*, **249**, 120–
2 137, doi:10.1016/J.AGRFORMET.2017.11.022.
3 <https://www.sciencedirect.com/science/article/pii/S0168192317303957> (Accessed May 17,
4 2018).
- 5 Oliver, D. M., R. D. Fish, M. Winter, C. J. Hodgson, A. L. Heathwaite, and D. R. Chadwick, 2012:
6 Valuing local knowledge as a source of expert data: Farmer engagement and the design of
7 decision support systems. *Environ. Model. Softw.*, **36**, 76–85,
8 doi:10.1016/J.ENVSOF.2011.09.013.
9 <https://www.sciencedirect.com/science/article/pii/S1364815211002118> (Accessed October 28,
10 2018).
- 11 Olsson, L., and Coauthors, 2014a: Cross-chapter box on heat stress and heat waves. *Climate Change*
12 *2014: Impacts, Adaptation, and Vulnerability. Part A: Global and Sectoral Aspects.*
13 *Contribution of Working Group II to the Fifth Assessment Report of the Intergovernmental*
14 *Panel on Climate Change*, K.J. Field, C.B., V.R. Barros, D.J. Dokken, A. Mach, M.D.
15 Mastrandrea, T.E. Bilir, M. Chatterjee, K.L. Ebi, Y.O. Estrada, R.C. Genova, B. Girma, E.S.
16 Kissel, A.N. Levy, S. MacCracken, P.R. Mastrandrea, and L.L. White, Eds., Cambridge
17 University Press, Cambridge, UK and New York, USA, 109–111.
- 18 Olsson, L., M. Opondo, P. Tschakert, A. Agrawal, S. Eriksen, S. Ma, L. Perch, and S. Zakieldeen,
19 2014b: Livelihoods and Poverty. *Climate change 2014 : impacts, adaptation, and vulnerability :*
20 *Working Group II contribution to the fifth assessment report of the Intergovernmental Panel on*
21 *Climate Change*, C.B. Field, V.R. Barros, and Et al, Eds., Cambridge University Press,
22 Cambridge, UK, and New York, USA, 793–832
23 <https://brage.bibsys.no/xmlui/handle/11250/2387830> (Accessed December 9, 2017).
- 24 Olsson, L., A. Jerneck, H. Thoren, J. Persson, and D. O’Byrne, 2015: Why resilience is unappealing
25 to social science: Theoretical and empirical investigations of the scientific use of resilience. *Sci.*
26 *Adv.*, **1**, e1400217–e1400217, doi:10.1126/sciadv.1400217.
27 <http://advances.sciencemag.org/cgi/doi/10.1126/sciadv.1400217> (Accessed May 25, 2018).
- 28 Omondi, M. O., X. Xia, A. Nahayo, X. Liu, P. K. Korai, and G. Pan, 2016: Quantification of biochar
29 effects on soil hydrological properties using meta-analysis of literature data. *Geoderma*, **274**,
30 28–34, doi:10.1016/J.GEODERMA.2016.03.029.
31 <https://www.sciencedirect.com/science/article/pii/S0016706116301471> (Accessed December 2,
32 2017).
- 33 Orr, B. J., and Coauthors, 2017a: Scientific conceptual framework for land degradation neutrality. *A*
34 *Report of the Science-Policy Interface. United Nations Convention to Combat Desertification*
35 *(UNCCD), Bonn, Germany.*
- 36 ———, and Coauthors, 2017b: *Scientific conceptual framework for land degradation neutrality.* Bonn,
37 Germany,.
- 38 Osborne, T. M., 2011: Carbon forestry and agrarian change: access and land control in a Mexican
39 rainforest. *J. Peasant Stud.*, **38**, 859–883, doi:10.1080/03066150.2011.611281.
40 <http://www.tandfonline.com/doi/abs/10.1080/03066150.2011.611281> (Accessed April 12,
41 2019).
- 42 Osland, M. J., N. M. Enwright, R. H. Day, C. A. Gabler, C. L. Stagg, and J. B. Grace, 2016: Beyond
43 just sea-level rise: considering macroclimatic drivers within coastal wetland vulnerability
44 assessments to climate change. *Glob. Chang. Biol.*, **22**, 1–11, doi:10.1111/gcb.13084.
45 <http://doi.wiley.com/10.1111/gcb.13084> (Accessed March 18, 2019).
- 46 Ostrom, E., 2009: A general framework for analyzing sustainability of social-ecological systems.
47 *Science*, **325**, 419–422, doi:10.1126/science.1172133.
48 <http://www.ncbi.nlm.nih.gov/pubmed/19628857> (Accessed November 2, 2018).
- 49 Ovalle-Rivera, O., P. Läderach, C. Bunn, M. Obersteiner, and G. Schroth, 2015: Projected Shifts in

- 1 Coffea arabica Suitability among Major Global Producing Regions Due to Climate Change.
2 *PLoS One*, **10**, e0124155, doi:10.1371/journal.pone.0124155.
3 <http://dx.plos.org/10.1371/journal.pone.0124155> (Accessed May 22, 2018).
- 4 Owour, B., W. Mauta, and S. Eriksen, 2011: Sustainable adaptation and human security: Interactions
5 between pastoral and agropastoral groups in dryland Kenya. *Clim. Dev.*, **3**, 42–58,
6 doi:10.3763/cdev.2010.0063. <http://www.tandfonline.com/doi/abs/10.3763/cdev.2010.0063>
7 (Accessed May 16, 2018).
- 8 Paetsch, L., C. W. Mueller, I. Kögel-Knabner, M. Von Lützow, C. Girardin, and C. Rumpel, 2018:
9 Effect of in-situ aged and fresh biochar on soil hydraulic conditions and microbial C use under
10 drought conditions. *Sci. Rep.*, doi:10.1038/s41598-018-25039-x.
- 11 Page, S. E., F. Siegert, J. O. Rieley, H.-D. V. Boehm, A. Jaya, and S. Limin, 2002: The amount of
12 carbon released from peat and forest fires in Indonesia during 1997. *Nature*, **420**, 61–65,
13 doi:10.1038/nature01131. <http://www.nature.com/articles/nature01131> (Accessed November 1,
14 2018).
- 15 Paladino, S., and S. J. Fiske, 2017: *The carbon fix : forest carbon, social justice, and environmental*
16 *governance*. Routledge, New York, United States, 320 pp.
- 17 Palm, C., H. Blanco-Canqui, F. DeClerck, L. Gatere, and P. Grace, 2014: Conservation agriculture
18 and ecosystem services: An overview. *Agric. Ecosyst. Environ.*, **187**, 87–105,
19 doi:10.1016/J.AGEE.2013.10.010.
20 <https://www.sciencedirect.com/science/article/pii/S0167880913003502> (Accessed February 21,
21 2019).
- 22 Pan, Y., and Coauthors, 2011: A large and persistent carbon sink in the world's forests. *Science* (80-
23), **333**, 988–993, doi:10.1126/science.1201609.
- 24 Panagos, P., P. Borrelli, K. Meusburger, C. Alewell, E. Lugato, and L. Montanarella, 2015:
25 Estimating the soil erosion cover-management factor at the European scale. *Land use policy*, **48**,
26 38–50, doi:10.1016/J.LANDUSEPOL.2015.05.021.
27 <https://www.sciencedirect.com/science/article/pii/S0264837715001611> (Accessed April 3,
28 2019).
- 29 Panfil, S. N., and C. A. Harvey, 2016: REDD+ and Biodiversity Conservation: A Review of the
30 Biodiversity Goals, Monitoring Methods, and Impacts of 80 REDD+ Projects. *Conserv. Lett.*, **9**,
31 143–150, doi:10.1111/conl.12188. <http://doi.wiley.com/10.1111/conl.12188> (Accessed
32 November 2, 2018).
- 33 Panthou, G., T. Vischel, and T. Lebel, 2014: Recent trends in the regime of extreme rainfall in the
34 Central Sahel. *Int. J. Climatol.*, **34**, 3998–4006, doi:10.1002/joc.3984.
35 <http://doi.wiley.com/10.1002/joc.3984> (Accessed October 1, 2018).
- 36 Parajuli, P. B., P. Jayakody, G. F. Sassenrath, and Y. Ouyang, 2016: Assessing the impacts of climate
37 change and tillage practices on stream flow, crop and sediment yields from the Mississippi River
38 Basin. *Agric. Water Manag.*, **168**, 112–124, doi:10.1016/J.AGWAT.2016.02.005.
39 <https://www.sciencedirect.com/science/article/pii/S0378377416300427> (Accessed May 23,
40 2018).
- 41 Pardini, R., A. de A. Bueno, T. A. Gardner, P. I. Prado, and J. P. Metzger, 2010: Beyond the
42 Fragmentation Threshold Hypothesis: Regime Shifts in Biodiversity Across Fragmented
43 Landscapes. *PLoS One*, **5**, e13666, doi:10.1371/journal.pone.0013666.
44 <http://dx.plos.org/10.1371/journal.pone.0013666> (Accessed October 1, 2018).
- 45 Parry, L. E., J. Holden, and P. J. Chapman, 2014: Restoration of blanket peatlands. *J. Environ.*
46 *Manage.*, doi:10.1016/j.jenvman.2013.11.033.
- 47 Pataki, D. E., and Coauthors, 2011: Coupling biogeochemical cycles in urban environments:
48 ecosystem services, green solutions, and misconceptions. *Front. Ecol. Environ.*, **9**, 27–36,
49 doi:10.1890/090220. <http://doi.wiley.com/10.1890/090220> (Accessed May 17, 2018).

- 1 Patange, O. S., and Coauthors, 2015: Reductions in Indoor Black Carbon Concentrations from
2 Improved Biomass Stoves in Rural India. *Environ. Sci. Technol.*, **49**, 4749–4756,
3 doi:10.1021/es506208x. <https://doi.org/10.1021/es506208x>.
- 4 Pattanayak, S. K., D. Evan Mercer, E. Sills, and J.-C. Yang, 2003: Taking stock of agroforestry
5 adoption studies. *Agrofor. Syst.*, **57**, 173–186, doi:10.1023/A:1024809108210.
6 <http://link.springer.com/10.1023/A:1024809108210> (Accessed October 22, 2018).
- 7 Paustian, K., J. Lehmann, S. Ogle, D. Reay, G. P. Robertson, and P. Smith, 2016: Climate-smart soils.
8 *Nature*, **532**, 49–57, doi:10.1038/nature17174.
9 <http://www.nature.com/doi/10.1038/nature17174> (Accessed May 21, 2018).
- 10 Payo, A., and Coauthors, 2016: Projected changes in area of the Sundarban mangrove forest in
11 Bangladesh due to SLR by 2100. *Clim. Change*, **139**, 279–291, doi:10.1007/s10584-016-1769-z.
12 <http://link.springer.com/10.1007/s10584-016-1769-z> (Accessed June 19, 2018).
- 13 Pearson, T. R. H., S. Brown, and F. M. Casarim, 2014: Carbon emissions from tropical forest
14 degradation caused by logging. *Environ. Res. Lett.*, **9**, 034017, doi:10.1088/1748-
15 9326/9/3/034017. [http://stacks.iop.org/1748-
16 9326/9/i=3/a=034017?key=crossref.c4aa12693bdced1aa5b0b6cf0316f44d](http://stacks.iop.org/1748-9326/9/i=3/a=034017?key=crossref.c4aa12693bdced1aa5b0b6cf0316f44d) (Accessed October 3,
17 2018).
- 18 Pearson, T. R. H., S. Brown, L. Murray, and G. Sidman, 2017: Greenhouse gas emissions from
19 tropical forest degradation: an underestimated source. *Carbon Balance Manag.*, **12**, 3,
20 doi:10.1186/s13021-017-0072-2. [http://cbmjournal.springeropen.com/articles/10.1186/s13021-
21 017-0072-2](http://cbmjournal.springeropen.com/articles/10.1186/s13021-017-0072-2) (Accessed May 26, 2018).
- 22 Pedlar, J. H., and Coauthors, 2012: Placing Forestry in the Assisted Migration Debate. *Bioscience*, **62**,
23 835–842, doi:10.1525/bio.2012.62.9.10. <https://doi.org/10.1525/bio.2012.62.9.10>.
- 24 Pellegrini, A. F. A., and Coauthors, 2018: Fire frequency drives decadal changes in soil carbon and
25 nitrogen and ecosystem productivity. doi:10.1038/nature24668.
26 <https://www.nature.com/articles/nature24668.pdf?origin=ppub> (Accessed October 25, 2018).
- 27 Van Pelt, R. S., S. X. Hushmurodov, R. L. Baumhardt, A. Chappell, M. A. Nearing, V. O. Polyakov,
28 and J. E. Strack, 2017: The reduction of partitioned wind and water erosion by conservation
29 agriculture. *CATENA*, **148**, 160–167, doi:10.1016/J.CATENA.2016.07.004.
30 <https://www.sciencedirect.com/science/article/pii/S0341816216302636> (Accessed April 3,
31 2019).
- 32 Peltzer, D. A., A. R. B., G. M. Lovett, D. Whitehead, and D. A. Wardle, 2010: Effects of biological
33 invasions on forest carbon sequestration. *Glob. Chang. Biol.*, **16**, 732–746, doi:10.1111/j.1365-
34 2486.2009.02038.x. <http://doi.wiley.com/10.1111/j.1365-2486.2009.02038.x> (Accessed
35 February 28, 2019).
- 36 Pendergrass, A. G., 2018: What precipitation is extreme? *Science (80-.)*, **360**, 1072–1073,
37 doi:10.1126/science.aat1871.
- 38 —, and R. Knutti, 2018: The Uneven Nature of Daily Precipitation and Its Change. *Geophys. Res.*
39 *Let.*, **45**, 11,980–11,988, doi:10.1029/2018GL080298.
40 <http://doi.wiley.com/10.1029/2018GL080298> (Accessed April 18, 2019).
- 41 —, —, F. Lehner, C. Deser, and B. M. Sanderson, 2017: Precipitation variability increases in a
42 warmer climate. *Sci. Rep.*, **7**, 17966, doi:10.1038/s41598-017-17966-y.
43 <http://www.nature.com/articles/s41598-017-17966-y> (Accessed April 18, 2019).
- 44 Pendleton, L., and Coauthors, 2012: Estimating Global “Blue Carbon” Emissions from Conversion
45 and Degradation of Vegetated Coastal Ecosystems. *PLoS One*, **7**, e43542,
46 doi:10.1371/journal.pone.0043542. <http://www.ncbi.nlm.nih.gov/pubmed/22962585> (Accessed
47 May 12, 2018).
- 48 Peng, X., and Coauthors, 2016: Response of changes in seasonal soil freeze/thaw state to climate

- 1 change from 1950 to 2010 across china. *J. Geophys. Res. Earth Surf.*, **121**, 1984–2000,
2 doi:10.1002/2016JF003876. <http://westdc.westgis.ac.cn/> (Accessed March 16, 2019).
- 3 Peng, X., Y. Deng, Y. Peng, and K. Yue, 2018: Effects of biochar addition on toxic element
4 concentrations in plants: A meta-analysis. *Sci. Total Environ.*, **616–617**, 970–977.
- 5 Percival, V., and T. Homer-Dixon, 1995: Environmental scarcity and violent conflict: the case of
6 Rwanda. <https://www.popline.org/node/296916> (Accessed May 15, 2018).
- 7 Peres, C. A., T. Emilio, J. Schiatti, S. J. M. Desmoulière, and T. Levi, 2016: Dispersal limitation
8 induces long-term biomass collapse in overhunted Amazonian forests. *Proc. Natl. Acad. Sci. U.*
9 *S. A.*, **113**, 892–897, doi:10.1073/pnas.1516525113.
10 <http://www.ncbi.nlm.nih.gov/pubmed/26811455> (Accessed November 2, 2018).
- 11 Perugini, L., L. Caporaso, S. Marconi, A. Cescatti, B. Quesada, N. de Noblet-Ducoudré, J. I. House,
12 and A. Arneth, 2017: Biophysical effects on temperature and precipitation due to land cover
13 change. *Environ. Res. Lett.*, **12**, 053002, doi:10.1088/1748-9326/aa6b3f.
14 [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/12/i=5/a=053002?key=crossref.c8fcedc7f3570d3a8b7baf3fe79f149c)
15 [9326/12/i=5/a=053002?key=crossref.c8fcedc7f3570d3a8b7baf3fe79f149c](http://stacks.iop.org/1748-9326/12/i=5/a=053002?key=crossref.c8fcedc7f3570d3a8b7baf3fe79f149c) (Accessed December
16 26, 2017).
- 17 Peters, P. E., 2004: Inequality and Social Conflict Over Land in Africa. *J. Agrar. Chang.*, **4**, 269–314,
18 doi:10.1111/j.1471-0366.2004.00080.x. <http://doi.wiley.com/10.1111/j.1471-0366.2004.00080.x>
19 (Accessed April 16, 2018).
- 20 Petzold, J., 2016: Limitations and opportunities of social capital for adaptation to climate change: a
21 case study on the Isles of Scilly. *Geogr. J.*, **182**, 123–134, doi:10.1111/geoj.12154.
- 22 —, and A. K. Magnan, 2019: Climate change: thinking small islands beyond Small Island
23 Developing States (SIDS). *Clim. Change*, **152**, 145–165, doi:10.1007/s10584-018-2363-3.
24 <http://link.springer.com/10.1007/s10584-018-2363-3> (Accessed April 10, 2019).
- 25 Phelps, J., E. L. Webb, and A. Agrawal, 2010: Does REDD+ threaten to recentralize forest
26 governance? *Science (80-.)*, **328**, 312–313, doi:10.1126/science.1187774.
27 <http://www.sciencemag.org/cgi/doi/10.1126/science.1187774> (Accessed October 25, 2018).
- 28 Pielke, R. A., J. Adegoke, A. Beltraán-Przekurat, C. A. Hiemstra, J. Lin, U. S. Nair, D. Niyogi, and
29 T. E. Nobis, 2007: An overview of regional land-use and land-cover impacts on rainfall. *Tellus*
30 *B Chem. Phys. Meteorol.*, **59**, 587–601, doi:10.1111/j.1600-0889.2007.00251.x.
31 <https://www.tandfonline.com/doi/full/10.1111/j.1600-0889.2007.00251.x> (Accessed October 22,
32 2018).
- 33 Piguet, E., R. Kaenzig, and J. Guélat, 2018: The uneven geography of research on “environmental
34 migration.” *Popul. Environ.*, 1–27, doi:10.1007/s11111-018-0296-4.
35 <http://link.springer.com/10.1007/s11111-018-0296-4> (Accessed May 25, 2018).
- 36 Piñeiro, G., J. M. Paruelo, M. Oesterheld, and E. G. Jobbágy, 2010: Pathways of Grazing Effects on
37 Soil Organic Carbon and Nitrogen. *Rangel. Ecol. Manag.*, **63**, 109–119, doi:10.2111/08-255.1.
38 <https://www.sciencedirect.com/science/article/pii/S1550742410500083> (Accessed March 15,
39 2019).
- 40 Pingoud, K., T. Ekholm, R. Sievänen, S. Huuskonen, and J. Hynynen, 2018: Trade-offs between
41 forest carbon stocks and harvests in a steady state – A multi-criteria analysis. *J. Environ.*
42 *Manage.*, **210**, 96–103, doi:10.1016/j.jenvman.2017.12.076.
- 43 Pinty, B., J.-L. Widlowski, M. M. Verstraete, I. Andreadakis, O. Arino, M. Clerici, T. Kaminski, and
44 M. Taberner, 2011: Snowy backgrounds enhance the absorption of visible light in forest
45 canopies. doi:10.1029/2010GL046417. <https://lpdaac.usgs.gov/> (Accessed March 16, 2019).
- 46 Piovano, E. L., D. Ariztegui, S. M. Bernasconi, and J. A. McKenzie, 2004: Stable isotopic record of
47 hydrological changes in subtropical Laguna Mar Chiquita (Argentina) over the last 230 years.
48 *The Holocene*, **14**, 525–535, doi:10.1191/0959683604hl729rp.

- 1 <http://journals.sagepub.com/doi/10.1191/0959683604hl729rp> (Accessed March 18, 2019).
- 2 Plangoen, P., and M. S. Babel, 2014: Projected rainfall erosivity changes under future climate in the
3 Upper Nan Watershed, Thailand. *J. Earth Sci. Clim. Change*, **5**.
- 4 Planque, C., D. Carrer, and J.-L. Roujean, 2017: Analysis of MODIS albedo changes over steady
5 woody covers in France during the period of 2001–2013. *Remote Sens. Environ.*, **191**, 13–29,
6 doi:10.1016/J.RSE.2016.12.019.
7 <https://www.sciencedirect.com/science/article/pii/S0034425716304989> (Accessed October 3,
8 2018).
- 9 Poeplau, C., and A. Don, 2015: Carbon sequestration in agricultural soils via cultivation of cover
10 crops – A meta-analysis. *Agric. Ecosyst. Environ.*, **200**, 33–41,
11 doi:10.1016/J.AGEE.2014.10.024.
12 <https://www.sciencedirect.com/science/article/pii/S0167880914004873> (Accessed May 11,
13 2018).
- 14 Poesen, J., J. Nachtergaele, G. Verstraeten, and C. Valentin, 2003: Gully erosion and environmental
15 change: importance and research needs. *CATENA*, **50**, 91–133, doi:10.1016/S0341-
16 8162(02)00143-1. <https://www.sciencedirect.com/science/article/pii/S0341816202001431>
17 (Accessed November 2, 2018).
- 18 Poesen, J. W. A., and J. M. Hooke, 1997: Erosion, flooding and channel management in
19 Mediterranean environments of southern Europe. *Prog. Phys. Geogr.*, **21**, 157–199,
20 doi:10.1177/030913339702100201.
21 <http://journals.sagepub.com/doi/10.1177/030913339702100201> (Accessed May 21, 2018).
- 22 Poff, N. L., 2002: Ecological response to and management of increased flooding caused by climate
23 change. *Philos. Trans. R. Soc. A Math. Phys. Eng. Sci.*, **360**, 1497–1510,
24 doi:10.1098/rsta.2002.1012.
25 <http://rsta.royalsocietypublishing.org/cgi/doi/10.1098/rsta.2002.1012> (Accessed December 27,
26 2017).
- 27 Pokam, W. M., C. L. Bain, R. S. Chadwick, R. Graham, D. J. Sonwa, and F. M. Kamga, 2014:
28 Identification of Processes Driving Low-Level Westerlies in West Equatorial Africa. *J. Clim.*,
29 **27**, 4245–4262, doi:10.1175/JCLI-D-13-00490.1. [http://journals.ametsoc.org/doi/10.1175/JCLI-
30 D-13-00490.1](http://journals.ametsoc.org/doi/10.1175/JCLI-D-13-00490.1) (Accessed November 1, 2018).
- 31 Polley, H. W., D. D. Briske, J. A. Morgan, K. Wolter, D. W. Bailey, and J. R. Brown, 2013: Climate
32 Change and North American Rangelands: Trends, Projections, and Implications. *Rangel. Ecol.*
33 *Manag.*, **66**, 493–511, doi:10.2111/REM-D-12-00068.1.
34 <http://linkinghub.elsevier.com/retrieve/pii/S1550742413500595> (Accessed March 6, 2018).
- 35 Pontee, N., 2013: Defining coastal squeeze: A discussion. *Ocean Coast. Manag.*, **84**, 204–207,
36 doi:10.1016/J.OCECOAMAN.2013.07.010.
37 <https://www.sciencedirect.com/science/article/pii/S0964569113001786> (Accessed October 27,
38 2018).
- 39 Poorter, L., and Coauthors, 2016: Biomass resilience of Neotropical secondary forests. *Nature*, **530**,
40 211–214. <https://doi.org/10.1038/nature16512>.
- 41 Porter-Bolland, L., E. A. Ellis, M. R. Guariguata, I. Ruiz-Mallén, S. Negrete-Yankelevich, and V.
42 Reyes-García, 2012: Community managed forests and forest protected areas: An assessment of
43 their conservation effectiveness across the tropics. *For. Ecol. Manage.*, **268**, 6–17,
44 doi:10.1016/J.FORECO.2011.05.034.
45 <https://www.sciencedirect.com/science/article/pii/S0378112711003215> (Accessed October 28,
46 2018).
- 47 Porter, J., and Coauthors, 2014: Climate change 2014: impacts, adaptation, and vulnerability:
48 Working Group II contribution to the fifth assessment report of the Intergovernmental Panel on
49 Climate Change. *Climate change 2014: impacts, adaptation, and vulnerability. Part A: global*

- 1 *and sectoral aspects*, Christopher B. Field et al., Eds., Cambridge University Press, New York,
2 United States, 485–533 <https://espace.library.uq.edu.au/view/UQ:353459> (Accessed March 15,
3 2018).
- 4 Porter, J. H., M. L. Parry, and T. R. Carter, 1991: The potential effects of climatic change on
5 agricultural insect pests. *Agric. For. Meteorol.*, **57**, 221–240, doi:10.1016/0168-1923(91)90088-
6 8. <https://www.sciencedirect.com/science/article/pii/0168192391900888> (Accessed October 24,
7 2018).
- 8 Post, W. M., and K. C. Kwon, 2000: Soil carbon sequestration and land-use change: processes and
9 potential. *Glob. Chang. Biol.*, **6**, 317–327, doi:10.1046/j.1365-2486.2000.00308.x.
10 <http://doi.wiley.com/10.1046/j.1365-2486.2000.00308.x> (Accessed May 21, 2018).
- 11 Potapov, P., and Coauthors, 2008: Mapping the World’s Intact Forest Landscapes by Remote Sensing.
12 *Ecol. Soc.*, **13**, art51, doi:10.5751/ES-02670-130251.
13 <http://www.ecologyandsociety.org/vol13/iss2/art51/> (Accessed May 26, 2018).
- 14 Potschin, M., R. H. (Roy H. . Haines-Young, R. Fish, and R. K. Turner, 2016: *Routledge handbook of*
15 *ecosystem services*. 629 pp.
- 16 Poulton, P., J. Johnston, A. Macdonald, R. White, and D. Powlson, 2018: Major limitations to
17 achieving “4 per 1000” increases in soil organic carbon stock in temperate regions: Evidence
18 from long-term experiments at Rothamsted Research, United Kingdom. *Glob. Chang. Biol.*, **24**,
19 2563–2584, doi:10.1111/gcb.14066. <http://doi.wiley.com/10.1111/gcb.14066> (Accessed March
20 29, 2019).
- 21 Powell, B., J. Hall, and T. Johns, 2011: Forest cover, use and dietary intake in the East Usambara
22 Mountains, Tanzania. *Int. For. Rev.*, **13**, 305–317, doi:10.1505/146554811798293944.
23 <http://www.ingentaconnect.com/content/10.1505/146554811798293944> (Accessed March 8,
24 2019).
- 25 Powlson, D. S., C. M. Stirling, M. L. Jat, B. G. Gerard, C. A. Palm, P. A. Sanchez, and K. G.
26 Cassman, 2014: Limited potential of no-till agriculture for climate change mitigation. *Nat. Clim.*
27 *Chang.*, **4**, 678–683, doi:10.1038/nclimate2292. <http://www.nature.com/articles/nclimate2292>
28 (Accessed March 15, 2019).
- 29 Prein, A. F., C. Liu, K. Ikeda, S. B. Trier, R. M. Rasmussen, G. J. Holland, and M. P. Clark, 2017:
30 Increased rainfall volume from future convective storms in the US. *Nat. Clim. Chang.*, **7**, 880–
31 884, doi:10.1038/s41558-017-0007-7. <https://www.nature.com/articles/s41558-017-0007-7.pdf>
32 (Accessed March 16, 2018).
- 33 Preti, F., and N. Romano, 2014: Terraced landscapes: From an old best practice to a potential hazard
34 for soil degradation due to land abandonment. *Anthropocene*, **6**, 10–25,
35 doi:10.1016/J.ANCENE.2014.03.002.
36 <https://www.sciencedirect.com/science/article/pii/S2213305414000113> (Accessed October 28,
37 2018).
- 38 —, and Coauthors, 2018: Conceptualization of Water Flow Pathways in Agricultural Terraced
39 Landscapes. *L. Degrad. Dev.*, **29**, 651–662, doi:10.1002/ldr.2764.
40 <http://doi.wiley.com/10.1002/ldr.2764> (Accessed October 28, 2018).
- 41 Price, D. T., and Coauthors, 2013: Anticipating the consequences of climate change for Canada’s
42 boreal forest ecosystems. *Environ. Rev.*, **21**, 322–365, doi:10.1139/er-2013-0042.
- 43 Price, R., 2017: “Clean” Cooking Energy in Uganda – technologies, impacts, and key barriers and
44 enablers to market acceleration. Institute of Development Studies, Brighton, UK,
45 <https://opendocs.ids.ac.uk/opendocs/handle/123456789/13234> (Accessed April 10, 2019).
- 46 Prince, S., and Coauthors, 2018: Status and trends of land degradation and restoration and associated
47 changes in biodiversity and ecosystem functions. *The IPBES assessment report on land*
48 *degradation and restoration*, L. Montanarella, R. Scholes, and A. Brainich, Eds., Bonn,
49 Germanu, 221–338.

- 1 Prince, S. D., 2016: Where Does Desertification Occur? Mapping Dryland Degradation at Regional to
2 Global Scales. *The End of Desertification*, R. Behnke and M. Mortimore, Eds., Springer, Berlin,
3 Heidelberg, 225–263 http://link.springer.com/10.1007/978-3-642-16014-1_9 (Accessed May 25,
4 2018).
- 5 Prince, S. D., E. B. De Colstoun, and L. L. Kravitz, 1998: Evidence from rain-use efficiencies does
6 not indicate extensive Sahelian desertification. *Glob. Chang. Biol.*, **4**, 359–374,
7 doi:10.1046/j.1365-2486.1998.00158.x. <http://doi.wiley.com/10.1046/j.1365-2486.1998.00158.x>
8 (Accessed October 25, 2018).
- 9 Pritchard, S. G., 2011: Soil organisms and global climate change. *Plant Pathol.*, **60**, 82–99,
10 doi:10.1111/j.1365-3059.2010.02405.x. <http://doi.wiley.com/10.1111/j.1365-3059.2010.02405.x>
11 (Accessed March 18, 2019).
- 12 Pritchard, W. R., and Coauthors, 1992: *Assessment of animal agriculture in Sub-Saharan Africa*.
13 Morrilton : Winrock International, 169p. pp.
- 14 Pryor, S. C., and R. J. Barthelmie, 2010: Climate change impacts on wind energy: A review. *Renew.*
15 *Sustain. Energy Rev.*, **14**, 430–437, doi:10.1016/J.RSER.2009.07.028.
16 <https://www.sciencedirect.com/science/article/pii/S1364032109001713> (Accessed March 14,
17 2018).
- 18 Pulido, M., S. Schnabel, J. F. L. Contador, J. Lozano-Parra, and Á. Gómez-Gutiérrez, 2017: Selecting
19 indicators for assessing soil quality and degradation in rangelands of Extremadura (SW Spain).
20 *Ecol. Indic.*, **74**, 49–61, doi:10.1016/J.ECOLIND.2016.11.016.
21 <https://www.sciencedirect.com/science/article/pii/S1470160X16306537#bib0280> (Accessed
22 September 27, 2018).
- 23 Pureswaran, D. S., L. De Grandpré, D. Paré, A. Taylor, M. Barrette, H. Morin, J. Régnière, and D. D.
24 Kneeshaw, 2015: Climate-induced changes in host tree–insect phenology may drive ecological
25 state-shift in boreal forests. *Ecology*, **96**, 1480–1491, doi:10.1890/13-2366.1.
26 <http://doi.wiley.com/10.1890/13-2366.1> (Accessed March 18, 2019).
- 27 Putz, F. E., and Coauthors, 2012: Sustaining conservation values in selectively logged tropical forests:
28 the attained and the attainable. *Conserv. Lett.*, **5**, 296–303, doi:10.1111/j.1755-
29 263X.2012.00242.x. <http://doi.wiley.com/10.1111/j.1755-263X.2012.00242.x> (Accessed
30 October 31, 2018).
- 31 Qiu, B., G. Chen, Z. Tang, D. Lu, Z. Wang, and C. Chen, 2017: Assessing the Three-North Shelter
32 Forest Program in China by a novel framework for characterizing vegetation changes. *ISPRS J.*
33 *Photogramm. Remote Sens.*, **133**, 75–88, doi:10.1016/j.isprsjprs.2017.10.003.
34 <https://doi.org/10.1016/j.isprsjprs.2017.10.003>.
- 35 Le Quéré, C., and Coauthors, 2013: The global carbon budget 1959-2011. *Earth Syst. Sci. Data*, **5**,
36 165–185, doi:10.5194/essd-5-165-2013.
- 37 Quilliam, R. S., S. Rangecroft, B. A. Emmett, T. H. Deluca, and D. L. Jones, 2013: Is biochar a
38 source or sink for polycyclic aromatic hydrocarbon (PAH) compounds in agricultural soils?
39 *GCB Bioenergy*, **5**, 96–103, doi:10.1111/gcbb.12007. <http://doi.wiley.com/10.1111/gcbb.12007>
40 (Accessed April 2, 2019).
- 41 Quin, P. R., and Coauthors, 2014: Oil mallee biochar improves soil structural properties-A study with
42 x-ray micro-CT. *Agric. Ecosyst. Environ.*, doi:10.1016/j.agee.2014.03.022.
- 43 Quinlan, A. E., M. Berbés-Blázquez, L. J. Haider, and G. D. Peterson, 2016: Measuring and assessing
44 resilience: broadening understanding through multiple disciplinary perspectives. *J. Appl. Ecol.*,
45 **53**, 677–687, doi:10.1111/1365-2664.12550. <http://doi.wiley.com/10.1111/1365-2664.12550>
46 (Accessed October 29, 2018).
- 47 Quiñonero-Rubio, J. M., E. Nadeu, C. Boix-Fayos, and J. de Vente, 2016: Evaluation of the
48 Effectiveness of Forest Restoration and Check-Dams to Reduce Catchment Sediment Yield. *L.*
49 *Degrad. Dev.*, **27**, 1018–1031, doi:10.1002/ldr.2331. <http://doi.wiley.com/10.1002/ldr.2331>

- 1 (Accessed March 8, 2019).
- 2 Quinton, J. N., G. Govers, K. van Oost, R. D. Bardgett, K. Van Oost, and R. D. Bardgett, 2010: The
3 impact of agricultural soil erosion on biogeochemical cycling. *Nat. Geosci.*, **3**, 311–314,
4 doi:10.1038/ngeo838. <https://www.cabdirect.org/cabdirect/abstract/20103162956> (Accessed
5 May 7, 2018).
- 6 Qureshi, A. S., 2011: Water Management in the Indus Basin in Pakistan: Challenges and
7 Opportunities. *Mt. Res. Dev.*, **31**, 252–260, doi:10.1659/MRD-JOURNAL-D-11-00019.1.
8 <http://www.bioone.org/doi/10.1659/MRD-JOURNAL-D-11-00019.1> (Accessed April 14, 2019).
- 9 Rabalais, N. N., and Coauthors, 2014: Eutrophication-Driven Deoxygenation in the Coastal Ocean.
10 *Oceanography*, **27**, 172–183, doi:10.2307/24862133. <https://www.jstor.org/stable/24862133>
11 (Accessed March 20, 2019).
- 12 Rahmstorf, S., 2010: A new view on sea level rise. *Nat. Reports Clim. Chang.*, 44–45,
13 doi:10.1038/climate.2010.29. <http://www.nature.com/doi/10.1038/climate.2010.29>
14 (Accessed May 31, 2018).
- 15 Raleigh, C., and H. Urdal, 2007: Climate change, environmental degradation and armed conflict.
16 *Polit. Geogr.*, **26**, 674–694, doi:10.1016/J.POLGEO.2007.06.005.
17 <https://www.sciencedirect.com/science/article/pii/S096262980700087X> (Accessed October 19,
18 2017).
- 19 Ramankutty, N., J. A. Foley, J. Norman, and K. McSweeney, 2002: The global distribution of
20 cultivable lands: current patterns and sensitivity to possible climate change. *Glob. Ecol.*
21 *Biogeogr.*, **11**, 377–392, doi:10.1046/j.1466-822x.2002.00294.x.
22 <http://doi.wiley.com/10.1046/j.1466-822x.2002.00294.x> (Accessed May 22, 2018).
- 23 Ramchunder, S. J., L. E. Brown, and J. Holden, 2012: Catchment-scale peatland restoration benefits
24 stream ecosystem biodiversity. *J. Appl. Ecol.*, doi:10.1111/j.1365-2664.2011.02075.x.
- 25 Rametsteiner, E., and M. Simula, 2003: Forest certification - An instrument to promote sustainable
26 forest management? *J. Environ. Manage.*, **67**, 87–98, doi:10.1016/S0301-4797(02)00191-3.
- 27 Ramisch, J., J. Keeley, I. Scoones, and W. Wolmer, 2002: Crop-Livestock policy in Africa: What is to
28 be done? *Pathways of change in Africa: crops, livestock & livelihoods in Mali, Ethiopia &*
29 *Zimbabwe*, I. Scoones and W. Wolmer, Eds., James Currey Ltd.
- 30 Rasul, G., a Mahmood, a Sadiq, and S. I. Khan, 2012: Vulnerability of the Indus Delta to Climate
31 Change in Pakistan. *Pakistan J. Meteorol.*, **8**, 89–107.
- 32 Ratcliffe, S., and Coauthors, 2017: Biodiversity and ecosystem functioning relations in European
33 forests depend on environmental context. *Ecol. Lett.*, **20**, 1414–1426, doi:10.1111/ele.12849.
34 <http://doi.wiley.com/10.1111/ele.12849> (Accessed March 15, 2019).
- 35 Ratter, B. M. W., J. Petzold, and K. Sinane, 2016: Considering the locals: coastal construction and
36 destruction in times of climate change on Anjouan, Comoros. *Nat. Resour. Forum*, **40**, 112–126,
37 doi:10.1111/1477-8947.12102.
- 38 Ravi, S., D. D. Breshears, T. E. Huxman, and P. D’Odorico, 2010: Land degradation in drylands:
39 Interactions among hydrologic–aeolian erosion and vegetation dynamics. *Geomorphology*, **116**,
40 236–245, doi:10.1016/j.geomorph.2009.11.023.
41 <http://linkinghub.elsevier.com/retrieve/pii/S0169555X09005108> (Accessed December 27, 2017).
- 42 Reed, M. S., 2005: *Participatory rangeland monitoring and management in the Kalahari, Botswana*.
43 University of Leeds, <http://etheses.whiterose.ac.uk/376/> (Accessed May 16, 2018).
- 44 Reed, M. S., and L. . Stringer, 2016a: *Land degradation, desertification and climate change:
45 anticipating, assessing and adapting to future change*. New York, NY: Routledge.
- 46 Reed, M. S., and L. C. Stringer, 2016b: *Land Degradation, Desertification and Climate Change*.
47 Taylor and Francis, Milton Park, Abingdon, UK, 178 pp.

- 1 Reed, M. S., A. J. Dougill, and M. J. Taylor, 2007: Integrating local and scientific knowledge for
2 adaptation to land degradation: Kalahari rangeland management options. *L. Degrad. Dev.*, **18**,
3 249–268, doi:10.1002/ldr.777. <http://doi.wiley.com/10.1002/ldr.777> (Accessed April 16, 2018).
- 4 Reed, S. C., K. K. Coe, J. P. Sparks, D. C. Housman, T. J. Zelikova, and J. Belnap, 2012: Changes to
5 dryland rainfall result in rapid moss mortality and altered soil fertility. *Nat. Clim. Chang.*, **2**,
6 752–755, doi:10.1038/nclimate1596. <http://www.nature.com/articles/nclimate1596> (Accessed
7 October 2, 2018).
- 8 Reenberg, A., T. Birch-Thomsen, O. Mertz, B. Fog, and S. Christiansen, 2008: Adaptation of Human
9 Coping Strategies in a Small Island Society in the SW Pacific—50 Years of Change in the
10 Coupled Human–Environment System on Bellona, Solomon Islands. *Hum. Ecol.*, **36**, 807–819,
11 doi:10.1007/s10745-008-9199-9. <http://link.springer.com/10.1007/s10745-008-9199-9>
12 (Accessed May 14, 2018).
- 13 Regina, K., J. Sheehy, and M. Myllys, 2015: Mitigating greenhouse gas fluxes from cultivated organic
14 soils with raised water table. *Mitig. Adapt. Strateg. Glob. Chang.*, **20**, 1529–1544,
15 doi:10.1007/s11027-014-9559-2.
- 16 Reid, P., and C. Vogel, 2006: Living and responding to multiple stressors in South Africa—Glimpses
17 from KwaZulu-Natal. *Glob. Environ. Chang.*, **16**, 195–206,
18 doi:10.1016/J.GLOENVCHA.2006.01.003.
19 <https://www.sciencedirect.com/science/article/pii/S095937800600015X> (Accessed May 14,
20 2018).
- 21 Reinwarth, B., R. Petersen, and J. Baade, 2019: Inferring mean rates of sediment yield and catchment
22 erosion from reservoir siltation in the Kruger National Park, South Africa: An uncertainty
23 assessment. *Geomorphology*, **324**, 1–13, doi:10.1016/J.GEOMORPH.2018.09.007.
24 <https://www.sciencedirect.com/science/article/pii/S01695555X18303647> (Accessed March 8,
25 2019).
- 26 Reis, V., V. Hermoso, S. K. Hamilton, D. Ward, E. Fluet-Chouinard, B. Lehner, and S. Linke, 2017:
27 A Global Assessment of Inland Wetland Conservation Status. *Bioscience*, **67**, 523–533,
28 doi:10.1093/biosci/bix045. [https://academic.oup.com/bioscience/article-
29 lookup/doi/10.1093/biosci/bix045](https://academic.oup.com/bioscience/article-lookup/doi/10.1093/biosci/bix045) (Accessed April 16, 2019).
- 30 REN21, 2018: *Renewables 2018: Global status report*. Paris, France, 325 pp.
31 http://www.ren21.net/wp-content/uploads/2018/06/17-8652_GSR2018_FullReport_web_-1.pdf.
- 32 Rengasamy, P., 2006: World salinization with emphasis on Australia. *J. Exp. Bot.*, **57**, 1017–1023,
33 doi:10.1093/jxb/erj108. [http://academic.oup.com/jxb/article/57/5/1017/641287/World-
34 salinization-with-emphasis-on-Australia](http://academic.oup.com/jxb/article/57/5/1017/641287/World-salinization-with-emphasis-on-Australia) (Accessed March 8, 2019).
- 35 Renou-Wilson, F., G. Moser, D. Fallon, C. A. Farrell, C. Müller, and D. Wilson, 2018: Rewetting
36 degraded peatlands for climate and biodiversity benefits: Results from two raised bogs.
37 *Ecological Engineering*.
- 38 Reyer, C. P. O., and Coauthors, 2015: Forest resilience and tipping points at different spatio-temporal
39 scales: approaches and challenges. *J. Ecol.*, **103**, 5–15, doi:10.1111/1365-2745.12337.
40 <http://dx.doi.org/10.1111/1365-2745.12337>.
- 41 Richards, C., and K. Lyons, 2016: The new corporate enclosures: Plantation forestry, carbon markets
42 and the limits of financialised solutions to the climate crisis. *Land use policy*, **56**, 209–216,
43 doi:10.1016/J.LANDUSEPOL.2016.05.013.
44 <https://www.sciencedirect.com/science/article/pii/S0264837716304562> (Accessed April 12,
45 2019).
- 46 Rist, L., A. Felton, L. Samuelsson, C. Sandström, and O. Rosvall, 2013: A new paradigm for adaptive
47 management. *Ecol. Soc.*, **18**, doi:10.5751/ES-06183-180463.
- 48 Ritzema, H., S. Limin, K. Kusin, J. Jauhiainen, and H. Wösten, 2014: Canal blocking strategies for
49 hydrological restoration of degraded tropical peatlands in Central Kalimantan, Indonesia.

- 1 *Catena*, **114**, 11–20, doi:10.1016/j.catena.2013.10.009.
- 2 Rivera-Ferre, M. G., F. López-i-Gelats, M. Howden, P. Smith, J. F. Morton, and M. Herrero, 2016:
3 Re-framing the climate change debate in the livestock sector: mitigation and adaptation options.
4 *Wiley Interdiscip. Rev. Clim. Chang.*, **7**, 869–892, doi:10.1002/wcc.421.
- 5 Rizwan, M., S. Ali, M. F. Qayyum, M. Ibrahim, M. Zia-ur-Rehman, T. Abbas, and Y. S. Ok, 2016:
6 Mechanisms of biochar-mediated alleviation of toxicity of trace elements in plants: a critical
7 review. *Environ. Sci. Pollut. Res.*, **23**, 2230–2248, doi:10.1007/s11356-015-5697-7.
8 <https://doi.org/10.1007/s11356-015-5697-7>.
- 9 Roberts, D., 2010: Prioritizing climate change adaptation and local level resilience in Durban, South
10 Africa. *Environ. Urban.*, doi:10.1177/0956247810379948.
- 11 Roberts, D., and S. O’Donoghue, 2013: Urban environmental challenges and climate change action in
12 Durban, South Africa. *Environ. Urban.*, **25**, 299–319, doi:10.1177/0956247813500904.
13 <http://journals.sagepub.com/doi/10.1177/0956247813500904> (Accessed May 17, 2018).
- 14 Roberts, M. W., A. W. D’Amato, C. C. Kern, and B. J. Palik, 2016: Long-term impacts of variable
15 retention harvesting on ground-layer plant communities in *Pinus resinosa* forests. *J. Appl. Ecol.*,
16 **53**, 1106–1116, doi:10.1111/1365-2664.12656.
- 17 Robinson, N., R. J. Harper, and K. R. J. Smettem, 2006: Soil water depletion by *Eucalyptus* spp.
18 integrated into dryland agricultural systems. *Plant Soil*, **286**, 141–151, doi:10.1007/s11104-006-
19 9032-4. <http://link.springer.com/10.1007/s11104-006-9032-4> (Accessed April 11, 2019).
- 20 Rogelj, J., and Coauthors, 2018: Mitigation pathways compatible with 1.5°C in the context of
21 sustainable development. *Global Warming of 1.5 °C an IPCC special report on the impacts of*
22 *global warming of 1.5 °C above pre-industrial levels and related global greenhouse gas*
23 *emission pathways, in the context of strengthening the global response to the threat of climate*
24 *change*, Masson-Delmotte, S.C. V., P. Zhai, H.-O. Pörtner, D. Roberts, J. Skea, P.R. Shukla, A.
25 Pirani, W. Moufouma-Okia, C. Péan, R. Pidcock, and And T.W. J.B.R. Matthews, Y. Chen, X.
26 Zhou, M.I. Gomis, E. Lonnoy, T. Maycock, M. Tignor, Eds., 93–174
27 <http://www.ipcc.ch/report/sr15/>.
- 28 Roman-Cuesta, R. M., and Coauthors, 2016: Hotspots of gross emissions from the land use sector:
29 patterns, uncertainties, and leading emission sources for the period 2000–2005 in the tropics.
30 *Biogeosciences*, **13**, 4253–4269, doi:10.5194/bg-13-4253-2016.
- 31 Romero, C., and F. E. Putz, 2018: Theory-of-change development for the evaluation of forest
32 stewardship council certification of sustained timber yields from natural forests in Indonesia.
33 *Forests*, **9**, doi:10.3390/f9090547.
- 34 Romo-Leon, J. R., W. J. D. van Leeuwen, and A. Castellanos-Villegas, 2014: Using remote sensing
35 tools to assess land use transitions in unsustainable arid agro-ecosystems. *J. Arid Environ.*, **106**,
36 27–35, doi:10.1016/j.jaridenv.2014.03.002.
- 37 Rosenzweig, C., and D. Hillel, 1998: *Climate change and the global harvest: potential impacts of the*
38 *greenhouse effect on agriculture*. Oxford University Press, Oxford, UK, 324 pp.
39 <https://www.cabdirect.org/cabdirect/abstract/19990707862> (Accessed May 15, 2018).
- 40 Rosenzweig, C., A. Iglesias, X. B. Yang, P. R. Epstein, and E. Chivian, 2001: Climate Change and
41 Extreme Weather Events; Implications for Food Production, Plant Diseases, and Pests. *Glob.*
42 *Chang. Hum. Heal.*, **2**, 90–104, doi:10.1023/A:1015086831467.
43 <http://link.springer.com/10.1023/A:1015086831467> (Accessed October 24, 2018).
- 44 —, and Coauthors, 2014: Assessing agricultural risks of climate change in the 21st century in a
45 global gridded crop model intercomparison. *Proc. Natl. Acad. Sci. U. S. A.*, **111**, 3268–3273,
46 doi:10.1073/pnas.1222463110. <http://www.ncbi.nlm.nih.gov/pubmed/24344314> (Accessed May
47 31, 2018).
- 48 Ross, N. J., 2011: Modern tree species composition reflects ancient Maya “forest gardens” in

- 1 northwest Belize. *Ecol. Appl.*, **21**, 75–84, doi:10.1890/09-0662.1.
2 <http://doi.wiley.com/10.1890/09-0662.1> (Accessed October 29, 2018).
- 3 Rossi, V., and Coauthors, 2017: Could REDD+ mechanisms induce logging companies to reduce
4 forest degradation in Central Africa? *J. For. Econ.*, **29**, 107–117,
5 doi:10.1016/J.JFE.2017.10.001.
6 <https://www.sciencedirect.com/science/article/pii/S1104689917301800> (Accessed October 31,
7 2018).
- 8 Rotenberg, E., and D. Yakir, 2010: Contribution of semi-arid forests to the climate system. *Science*,
9 **327**, 451–454, doi:10.1126/science.1179998. <http://www.ncbi.nlm.nih.gov/pubmed/20093470>
10 (Accessed October 31, 2018).
- 11 Routschek, A., J. Schmidt, and F. Kreienkamp, 2014: Impact of climate change on soil erosion — A
12 high-resolution projection on catchment scale until 2100 in Saxony/Germany. *CATENA*, **121**,
13 99–109, doi:10.1016/J.CATENA.2014.04.019.
14 <https://www.sciencedirect.com/science/article/pii/S0341816214001283> (Accessed October 28,
15 2017).
- 16 Royal Society, L., 2016: *Resilience to Extreme Weather*. Royal Society, London.
- 17 Rudel, T., and Coauthors, 2016: Do Smallholder, Mixed Crop-Livestock Livelihoods Encourage
18 Sustainable Agricultural Practices? A Meta-Analysis. *Land*, **5**, 6, doi:10.3390/land5010006.
19 <http://www.mdpi.com/2073-445X/5/1/6> (Accessed October 30, 2018).
- 20 Ruggiero, P., 2013: Is the Intensifying Wave Climate of the U.S. Pacific Northwest Increasing
21 Flooding and Erosion Risk Faster Than Sea-Level Rise? *J. Waterw. Port, Coastal, Ocean Eng.*,
22 **139**, 88–97, doi:10.1061/(ASCE)WW.1943-5460.0000172.
23 <http://ascelibrary.org/doi/10.1061/%28ASCE%29WW.1943-5460.0000172> (Accessed May 31,
24 2018).
- 25 Rumpel, C., F. Amiraslani, L.-S. Koutika, P. Smith, D. Whitehead, and E. Wollenberg, 2018: Put
26 more carbon in soils to meet Paris climate pledges. *Nature*, **564**, 32–34, doi:10.1038/d41586-
27 018-07587-4. <http://www.nature.com/articles/d41586-018-07587-4> (Accessed March 29, 2019).
- 28 Ruppert, J. C., A. Holm, S. Mieke, E. Muldavin, H. A. Snyman, K. Wesche, and A. Linstädter, 2012:
29 Meta-analysis of ANPP and rain-use efficiency confirms indicative value for degradation and
30 supports non-linear response along precipitation gradients in drylands. *J. Veg. Sci.*, **23**, 1035–
31 1050, doi:10.1111/j.1654-1103.2012.01420.x. [http://doi.wiley.com/10.1111/j.1654-
32 1103.2012.01420.x](http://doi.wiley.com/10.1111/j.1654-) (Accessed January 3, 2018).
- 33 Russell, M. B., S. Fraver, T. Aakala, J. H. Gove, C. W. Woodall, A. W. D’Amato, and M. J. Ducey,
34 2015: Quantifying carbon stores and decomposition in dead wood: A review. *For. Ecol.*
35 *Manage.*, **350**, 107–128, doi:10.1016/j.foreco.2015.04.033. [https://ac-els-cdn-
36 com.ezproxy.library.uvic.ca/S0378112715002558/1-s2.0-S0378112715002558-
37 main.pdf?_tid=b79add30-d563-11e7-8a2b-
38 00000aacb35f&acdnat=1512001172_93eebe3892e73dfc4cda94ac809eda6d](https://ac-els-cdn-com.ezproxy.library.uvic.ca/S0378112715002558/1-s2.0-S0378112715002558-main.pdf?_tid=b79add30-d563-11e7-8a2b-00000aacb35f&acdnat=1512001172_93eebe3892e73dfc4cda94ac809eda6d) (Accessed
39 November 29, 2017).
- 40 Rutherford, W. A., T. H. Painter, S. Ferrenberg, J. Belnap, G. S. Okin, C. Flagg, and S. C. Reed,
41 2017: Albedo feedbacks to future climate via climate change impacts on dryland biocrusts. *Sci.*
42 *Rep.*, **7**, 44188, doi:10.1038/srep44188. <http://www.nature.com/articles/srep44188> (Accessed
43 March 18, 2019).
- 44 Ryan, M. R., T. E. Crews, S. W. Culman, L. R. DeHaan, R. C. Hayes, J. M. Jungers, and M. G.
45 Bakker, 2018: Managing for Multifunctionality in Perennial Grain Crops. *Bioscience*, **68**, 294–
46 304, doi:10.1093/biosci/biy014. <https://academic.oup.com/bioscience/article/68/4/294/4942086>
47 (Accessed March 25, 2019).
- 48 Sachs, J. D., 2007: Poverty and environmental stress fuel Darfur crisis. *Nature*, **449**, 24–24,
49 doi:10.1038/449024a. <http://www.nature.com/doi/10.1038/449024a> (Accessed May 15,

- 1 2018).
- 2 Sainju, U. M., B. L. Allen, A. W. Lenssen, and R. P. Ghimire, 2017: Root biomass, root/shoot ratio,
3 and soil water content under perennial grasses with different nitrogen rates. *F. Crop. Res.*, **210**,
4 183–191, doi:10.1016/J.FCR.2017.05.029.
5 <https://www.sciencedirect.com/science/article/pii/S0378429017301132> (Accessed April 12,
6 2019).
- 7 Salehyan, I., 2008: From Climate Change to Conflict? No Consensus Yet. *J. Peace Res.*, **45**, 315–326,
8 doi:10.1177/0022343308088812. <http://journals.sagepub.com/doi/10.1177/0022343308088812>
9 (Accessed May 16, 2018).
- 10 Sanderman, J., T. Hengl, and G. J. Fiske, 2017: Soil carbon debt of 12,000 years of human land use.
11 *Proc. Natl. Acad. Sci. U. S. A.*, **114**, 9575–9580, doi:10.1073/pnas.1706103114.
12 <http://www.ncbi.nlm.nih.gov/pubmed/28827323> (Accessed March 27, 2019).
- 13 Sanquetta, C. R., A. P. Dalla Corte, A. L. Pelissari, M. Tomé, G. C. B. Maas, and M. N. I. Sanquetta,
14 2018: Dynamics of carbon and CO₂ removals by Brazilian forest plantations during 1990-2016.
15 *Carbon Balance Manag.*, **13**, doi:10.1186/s13021-018-0106-4. [https://doi.org/10.1186/s13021-](https://doi.org/10.1186/s13021-018-0106-4)
16 [018-0106-4](https://doi.org/10.1186/s13021-018-0106-4).
- 17 Santos, M. J., S. C. Dekker, V. Daioglou, M. C. Braakhekke, and D. P. van Vuuren, 2017: Modeling
18 the Effects of Future Growing Demand for Charcoal in the Tropics. *Front. Environ. Sci.*, **5**, 28.
19 <https://www.frontiersin.org/article/10.3389/fenvs.2017.00028>.
- 20 Sasaki, N., and F. E. Putz, 2009: Critical need for new definitions of “forest” and “forest degradation”
21 in global climate change agreements. *Conserv. Lett.*, **2**, 226–232, doi:10.1111/j.1755-
22 263X.2009.00067.x. <http://doi.wiley.com/10.1111/j.1755-263X.2009.00067.x>.
- 23 Sasaki, T., T. Okayasu, U. Jamsran, and K. Takeuchi, 2007: Threshold changes in vegetation along a
24 grazing gradient in Mongolian rangelands. *J. Ecol.*, **0**, 071106211313003–???,
25 doi:10.1111/j.1365-2745.2007.01315.x. <http://doi.wiley.com/10.1111/j.1365-2745.2007.01315.x>
26 (Accessed December 28, 2017).
- 27 Saugier, B., 2001: Estimations of Global Terrestrial Productivity: Converging Toward a Single
28 Number? *Terrestrial Global Productivity*, J. Roy, Ed., Academic Press, San Diego, CA, CA,
29 543–556.
- 30 Savard, J.-P., P. Bernatchez, F. Morneau, and F. Saucier, 2009: Vulnérabilité des communautés
31 côtières de l’est du Québec aux impacts des changements climatiques. *La Houille Blanche*, 59–
32 66, doi:10.1051/lhb/2009015. <http://www.shf-lhb.org/10.1051/lhb/2009015> (Accessed May 31,
33 2018).
- 34 Sayer, J., and Coauthors, 2013: Ten principles for a landscape approach to reconciling agriculture,
35 conservation, and other competing land uses. *Proc. Natl. Acad. Sci. U. S. A.*, **110**, 8349–8356,
36 doi:10.1073/pnas.1210595110. <http://www.ncbi.nlm.nih.gov/pubmed/23686581> (Accessed
37 March 15, 2019).
- 38 La Scala Júnior, N., E. De Figueiredo, and A. Panosso, 2012: A review on soil carbon accumulation
39 due to the management change of major Brazilian agricultural activities. *Brazilian J. Biol.*, **72**,
40 775–785, doi:10.1590/S1519-69842012000400012.
41 [http://www.scielo.br/scielo.php?script=sci_arttext&pid=S1519-](http://www.scielo.br/scielo.php?script=sci_arttext&pid=S1519-69842012000400012&lng=en&tlng=en)
42 [69842012000400012&lng=en&tlng=en](http://www.scielo.br/scielo.php?script=sci_arttext&pid=S1519-69842012000400012&lng=en&tlng=en) (Accessed October 18, 2018).
- 43 Scarano, F. R., 2017: Ecosystem-based adaptation to climate change: concept, scalability and a role
44 for conservation science. *Perspect. Ecol. Conserv.*, **15**, 65–73,
45 doi:10.1016/J.PECON.2017.05.003. [https://www.sciencedirect-](https://www.sciencedirect.com.ludwig.lub.lu.se/science/article/pii/S1679007316301621)
46 [com.ludwig.lub.lu.se/science/article/pii/S1679007316301621](https://www.sciencedirect.com.ludwig.lub.lu.se/science/article/pii/S1679007316301621) (Accessed March 28, 2019).
- 47 Schaefer, H., and Coauthors, 2016: A 21st-century shift from fossil-fuel to biogenic methane
48 emissions indicated by ¹³CH₄. *Science*, **352**, 80–84, doi:10.1126/science.aad2705.

- 1 <http://www.ncbi.nlm.nih.gov/pubmed/26966190> (Accessed October 3, 2018).
- 2 Scharlemann, J. P., E. V. Tanner, R. Hiederer, and V. Kapos, 2014: Global soil carbon: understanding
3 and managing the largest terrestrial carbon pool. *Carbon Manag.*, **5**, 81–91,
4 doi:10.4155/cmt.13.77. <http://www.tandfonline.com/doi/abs/10.4155/cmt.13.77> (Accessed
5 February 22, 2019).
- 6 Scheffers, B. R., and Coauthors, 2016: The broad footprint of climate change from genes to biomes to
7 people. *Science* (80-.), **354**, aaf7671, doi:10.1126/science.aaf7671.
8 <http://www.ncbi.nlm.nih.gov/pubmed/27846577> (Accessed March 18, 2019).
- 9 Scheffran, J., M. Brzoska, J. Kominek, P. M. Link, and J. Schilling, 2012: Disentangling the climate-
10 conflict nexus: empirical and theoretical assessment of vulnerabilities and pathways. *Rev. Eur.*
11 *Stud.*, **4**, 1. <http://www.ccsenet.org/journal/index.php/res/article/view/19277/13740> (Accessed
12 May 15, 2018).
- 13 Scheidel, A., and C. Work, 2018: Forest plantations and climate change discourses: New powers of
14 ‘green’ grabbing in Cambodia. *Land use policy*, **77**, 9–18,
15 doi:10.1016/J.LANDUSEPOL.2018.04.057.
16 <https://www.sciencedirect.com/science/article/abs/pii/S0264837717312401> (Accessed April 12,
17 2019).
- 18 Scherr, S. J., 2000: A downward spiral? Research evidence on the relationship between poverty and
19 natural resource degradation. *Food Policy*, **25**, 479–498, doi:10.1016/S0306-9192(00)00022-1.
20 <https://www.sciencedirect.com/science/article/pii/S0306919200000221> (Accessed May 15,
21 2018).
- 22 Schlautman, B., S. Barriball, C. Ciotir, S. Herron, and A. Miller, 2018: Perennial Grain Legume
23 Domestication Phase I: Criteria for Candidate Species Selection. *Sustainability*, **10**, 730,
24 doi:10.3390/su10030730. <http://www.mdpi.com/2071-1050/10/3/730> (Accessed May 21, 2018).
- 25 Schlesinger, W. H., 2009: On the fate of anthropogenic nitrogen. *Proc. Natl. Acad. Sci.*, **106**, 203 LP-
26 208. <http://www.pnas.org/content/106/1/203.abstract>.
- 27 Schlesinger, W. H., and S. Jasechko, 2014: Transpiration in the global water cycle. *Agric. For.*
28 *Meteorol.*, **189–190**, 115–117, doi:10.1016/J.AGRFORMET.2014.01.011.
29 <https://www.sciencedirect.com/science/article/pii/S0168192314000203#bib0060> (Accessed
30 March 5, 2018).
- 31 —, and R. Amundson, 2018: Managing for soil carbon sequestration: Let’s get realistic. *Glob.*
32 *Chang. Biol.*, **25**, gcb.14478, doi:10.1111/gcb.14478.
33 <https://onlinelibrary.wiley.com/doi/abs/10.1111/gcb.14478> (Accessed April 5, 2019).
- 34 Schleussner, C.-F., and Coauthors, 2016: Science and policy characteristics of the Paris Agreement
35 temperature goal. *Nat. Clim. Chang.*, **6**, 827–835, doi:10.1038/nclimate3096.
36 <http://www.nature.com/articles/nclimate3096> (Accessed March 15, 2018).
- 37 Schmidt, M. W. I., and Coauthors, 2011: Persistence of soil organic matter as an ecosystem property.
38 *Nature*, **478**, 49–56, doi:10.1038/nature10386.
39 <http://www.nature.com/doi/abs/10.1038/nature10386>.
- 40 Schnitzer, S. A., and Coauthors, 2011: Soil microbes drive the classic plant diversity–productivity
41 pattern. *Ecology*, **92**, 296–303.
- 42 Schoenholtz, S. ., H. V. Miegroet, and J. . Burger, 2000: A review of chemical and physical properties
43 as indicators of forest soil quality: challenges and opportunities. *For. Ecol. Manage.*, **138**, 335–
44 356, doi:10.1016/S0378-1127(00)00423-0.
45 <https://www.sciencedirect.com/science/article/pii/S0378112700004230> (Accessed March 18,
46 2019).
- 47 Schofield, R. V., and M. J. Kirkby, 2003: Application of salinization indicators and initial
48 development of potential global soil salinization scenario under climatic change. *Global*

- 1 *Biogeochem. Cycles*, **17**, n/a-n/a, doi:10.1029/2002GB001935.
2 <http://doi.wiley.com/10.1029/2002GB001935> (Accessed December 27, 2017).
- 3 Schuerch, M., and Coauthors, 2018a: Future response of global coastal wetlands to sea-level rise.
4 *Nature*, **561**, 231–234, doi:10.1038/s41586-018-0476-5. [http://www.nature.com/articles/s41586-](http://www.nature.com/articles/s41586-018-0476-5)
5 [018-0476-5](http://www.nature.com/articles/s41586-018-0476-5) (Accessed October 5, 2018).
- 6 ———, and Coauthors, 2018b: Future response of global coastal wetlands to sea-level rise. *Nature*, **561**,
7 231–234, doi:10.1038/s41586-018-0476-5. <http://www.nature.com/articles/s41586-018-0476-5>
8 (Accessed March 14, 2019).
- 9 Schut, A. G. T., E. Ivits, J. G. Conijn, B. ten Brink, and R. Fensholt, 2015: Trends in Global
10 Vegetation Activity and Climatic Drivers Indicate a Decoupled Response to Climate Change.
11 *PLoS One*, **10**, e0138013, doi:10.1371/journal.pone.0138013.
12 <https://dx.plos.org/10.1371/journal.pone.0138013> (Accessed February 20, 2019).
- 13 Schuur, E. A. G., and Coauthors, 2015: Climate change and the permafrost carbon feedback. *Nature*,
14 **520**, 171–179, doi:10.1038/nature14338. <http://www.nature.com/articles/nature14338> (Accessed
15 March 22, 2019).
- 16 Schwilch, G., and Coauthors, 2011: Experiences in monitoring and assessment of sustainable land
17 management. *L. Degrad. Dev.*, **22**, 214–225, doi:10.1002/ldr.1040.
18 <http://doi.wiley.com/10.1002/ldr.1040> (Accessed October 28, 2018).
- 19 Scoones, I., and W. Wolmer, 2002: Pathways of Change: Crop-Livestock Integration in Africa.
20 *Pathways of change in Africa: crops, livestock & livelihoods in Mali, Ethiopia & Zimbabwe*, I.
21 Scoones and W. Wolmer, Eds., James Currey Ltd.
- 22 Sealey, N. E., 2006: The cycle of casuarina-induced beach erosion - A case study from Andros,
23 Bahamas. *12th Symp. Geol. Bahamas other Carbonate Reg. San Salvador. Bahamas*, 196–204.
- 24 Seaquist, J. W., T. Hickler, L. Eklundh, J. Ardö, and B. W. Heumann, 2009: Disentangling the effects
25 of climate and people on Sahel vegetation dynamics. *Biogeosciences*, **6**, 469–477,
26 doi:10.5194/bg-6-469-2009. <http://www.biogeosciences.net/6/469/2009/> (Accessed October 10,
27 2018).
- 28 Sedano, F., and Coauthors, 2016: The impact of charcoal production on forest degradation: a case
29 study in Tete, Mozambique. *Environ. Res. Lett.*, **11**, 094020, doi:10.1088/1748-
30 9326/11/9/094020. [http://stacks.iop.org/1748-](http://stacks.iop.org/1748-9326/11/i=9/a=094020?key=crossref.a3ab07e751e2b171e389c7728e1c82fc)
31 [9326/11/i=9/a=094020?key=crossref.a3ab07e751e2b171e389c7728e1c82fc](http://stacks.iop.org/1748-9326/11/i=9/a=094020?key=crossref.a3ab07e751e2b171e389c7728e1c82fc) (Accessed May 23,
32 2018).
- 33 Segura, C., G. Sun, S. McNulty, and Y. Zhang, 2014: Potential impacts of climate change on soil
34 erosion vulnerability across the conterminous United States. *J. Soil Water Conserv.*, **69**, 171–
35 181, doi:10.2489/jswc.69.2.171. <http://www.jswconline.org/cgi/doi/10.2489/jswc.69.2.171>
36 (Accessed March 11, 2019).
- 37 Seidl, R., W. Rammer, D. Jäger, W. S. Currie, and M. J. Lexer, 2007: Assessing trade-offs between
38 carbon sequestration and timber production within a framework of multi-purpose forestry in
39 Austria. *For. Ecol. Manage.*, **248**, 64–79, doi:https://doi.org/10.1016/j.foreco.2007.02.035.
40 <http://www.sciencedirect.com/science/article/pii/S0378112707002101>.
- 41 ———, and Coauthors, 2017: Forest disturbances under climate change. *Nat. Clim. Chang.*, **7**, 395–402,
42 doi:10.1038/nclimate3303. <http://www.nature.com/articles/nclimate3303> (Accessed April 11,
43 2019).
- 44 Serpa, D., and Coauthors, 2015: Impacts of climate and land use changes on the hydrological and
45 erosion processes of two contrasting Mediterranean catchments. *Sci. Total Environ.*, **538**, 64–77,
46 doi:10.1016/J.SCITOTENV.2015.08.033.
47 <https://www.sciencedirect.com/science/article/pii/S0048969715305386?via%3Dihub> (Accessed
48 March 14, 2018).

- 1 Seto, K. C., and Coauthors, 2012: Urban land teleconnections and sustainability. *Proc. Natl. Acad. Sci.*, **109**, 7687–7692, doi:10.1073/pnas.1117622109.
- 2
- 3 Settele, J., and Coauthors, 2015: Terrestrial and Inland Water Systems. *Climate Change 2014*
- 4 *Impacts, Adaptation, and Vulnerability*, C.B. Field, V.R. Barros, D.J. Dokken, K.J. Mach, and
- 5 M.D. Mastrandrea, Eds., Cambridge University Press, Cambridge, 271–360
- 6 <http://ebooks.cambridge.org/ref/id/CBO9781107415379A025> (Accessed January 2, 2018).
- 7 Seymour, F., and A. Angelsen, 2012: Summary and Conclusions: REDD+ without regrets. *Analysing*
- 8 *REDD+: Challenges and choices*, Center for International Forestry Research (CIFOR), Bogor,
- 9 Indonesia, 317–334 <https://www.cifor.org/library/3832/> (Accessed October 31, 2018).
- 10 Shadkam, S., F. Ludwig, P. van Oel, İ. ½ağla Kirit, and P. Kabat, 2016: Impacts of climate change
- 11 and water resources development on the declining inflow into Iran’s Urmia Lake. *J. Great Lakes*
- 12 *Res.*, **42**, 942–952, doi:10.1016/j.jglr.2016.07.033.
- 13 Shakesby, R. A., 2011: Post-wildfire soil erosion in the Mediterranean: Review and future research
- 14 directions. *Earth-Science Rev.*, **105**, 71–100, doi:10.1016/J.EARSCIREV.2011.01.001.
- 15 <https://www.sciencedirect.com/science/article/pii/S001282521100002X> (Accessed March 18,
- 16 2019).
- 17 Shames, S., M. Hill Clarvis, and G. Kissinger, 2014: *Financing Strategies for Integrated Landscape*
- 18 *Investment*. Washington DC, USA, 1-60 pp.
- 19 Shanahan, T. M., and Coauthors, 2016: CO2 and fire influence tropical ecosystem stability in
- 20 response to climate change. *Sci. Rep.*, **6**, 29587, doi:10.1038/srep29587.
- 21 <http://www.nature.com/articles/srep29587> (Accessed November 1, 2018).
- 22 Shao, Y., 2008: *Physics and modelling of wind erosion*. Springer, 452 pp.
- 23 Sharmila, S., and K. J. E. Walsh, 2018: Recent poleward shift of tropical cyclone formation linked to
- 24 Hadley cell expansion. *Nat. Clim. Chang.*, **8**, 730–736, doi:10.1038/s41558-018-0227-5.
- 25 <http://www.nature.com/articles/s41558-018-0227-5> (Accessed October 14, 2018).
- 26 Sheffield, J., E. F. Wood, and M. L. Roderick, 2012: Little change in global drought over the past 60
- 27 years. *Nature*, **491**, 435–438, doi:10.1038/nature11575.
- 28 <http://www.nature.com/doi/10.1038/nature11575> (Accessed February 14, 2018).
- 29 Sheil, D., and D. Murdiyarso, 2009: How Forests Attract Rain: An Examination of a New Hypothesis.
- 30 *Bioscience*, **59**, 341–347, doi:10.1525/bio.2009.59.4.12.
- 31 <https://academic.oup.com/bioscience/article-lookup/doi/10.1525/bio.2009.59.4.12> (Accessed
- 32 November 1, 2018).
- 33 Shi, H., and Coauthors, 2017a: Assessing the ability of MODIS EVI to estimate terrestrial ecosystem
- 34 gross primary production of multiple land cover types. *Ecol. Indic.*, **72**, 153–164,
- 35 doi:10.1016/J.ECOLIND.2016.08.022.
- 36 <https://www.sciencedirect.com/science/article/pii/S1470160X16304836> (Accessed March 22,
- 37 2019).
- 38 —, and Coauthors, 2017b: Assessing the ability of MODIS EVI to estimate terrestrial ecosystem
- 39 gross primary production of multiple land cover types. *Ecol. Indic.*, **72**, 153–164,
- 40 doi:10.1016/j.ecolind.2016.08.022.
- 41 <https://www.sciencedirect.com/science/article/pii/S1470160X16304836> (Accessed March 18,
- 42 2019).
- 43 Shi, S., W. Zhang, P. Zhang, Y. Yu, and F. Ding, 2013: A synthesis of change in deep soil organic
- 44 carbon stores with afforestation of agricultural soils. *For. Ecol. Manage.*, **296**, 53–63,
- 45 doi:10.1016/j.foreco.2013.01.026. <http://dx.doi.org/10.1016/j.foreco.2013.01.026>.
- 46 Shindell, D., and Coauthors, 2012: Simultaneously Mitigating Near-Term Climate Change and
- 47 Improving Human Health and Food Security. *Science (80-.)*, **335**, 183 LP-189,
- 48 doi:10.1126/science.1210026. <http://science.sciencemag.org/content/335/6065/183.abstract>.

- 1 Shirzaei, M., and R. Bürgmann, 2018: Global climate change and local land subsidence exacerbate
2 inundation risk to the San Francisco Bay Area. *Sci. Adv.*, **4**, eaap9234,
3 doi:10.1126/sciadv.aap9234.
4 <http://advances.sciencemag.org/lookup/doi/10.1126/sciadv.aap9234> (Accessed March 17, 2019).
- 5 Showers, K. B., 2005: *Imperial gullies: soil erosion and conservation in Lesotho*. Ohio University
6 Press, 346 pp.
- 7 Shuab, R., R. Lone, J. Ahmad, and Z. A. Reshi, 2017: Arbuscular Mycorrhizal Fungi: A Potential
8 Tool for Restoration of Degraded Land. *Mycorrhiza - Nutrient Uptake, Biocontrol,*
9 *Ecorestoration*, Springer International Publishing, Cham, 415–434
10 http://link.springer.com/10.1007/978-3-319-68867-1_22 (Accessed March 6, 2019).
- 11 Siahaya, M. E., T. R. Hutaeruk, H. S. E. S. Aponno, J. W. Hatulesila, and A. B. Mardhanie, 2016:
12 Traditional ecological knowledge on shifting cultivation and forest management in East Borneo,
13 Indonesia. *Int. J. Biodivers. Sci. Ecosyst. Serv. Manag.*, **12**, 14–23,
14 doi:10.1080/21513732.2016.1169559.
15 <http://www.tandfonline.com/doi/full/10.1080/21513732.2016.1169559> (Accessed October 28,
16 2018).
- 17 Simon, J., A. Diego, P. Marija, B. Ana Catarina, G. Jan Willem van, A. H. Bruce, and V. Frank, 2017:
18 Biochar boosts tropical but not temperate crop yields. *Environ. Res. Lett.*, **12**, 53001.
19 <http://stacks.iop.org/1748-9326/12/i=5/a=053001>.
- 20 Simonsen, S. H., and Coauthors, 2014: *Applying resilience thinking: seven principles for building*
21 *resilience in social-ecological systems*.
22 [http://stockholmresilience.org/download/18.10119fc11455d3c557d6928/1459560241272/SRC+](http://stockholmresilience.org/download/18.10119fc11455d3c557d6928/1459560241272/SRC+Applying+Resilience+final.pdf)
23 [Applying+Resilience+final.pdf](http://stockholmresilience.org/download/18.10119fc11455d3c557d6928/1459560241272/SRC+Applying+Resilience+final.pdf).
- 24 Sims, N. C., J. R. England, G. J. Newnham, S. Alexander, C. Green, S. Minelli, and A. Held, 2019:
25 Developing good practice guidance for estimating land degradation in the context of the United
26 Nations Sustainable Development Goals. *Environ. Sci. Policy*, **92**, 349–355,
27 doi:10.1016/J.ENVSCI.2018.10.014.
28 <https://www.sciencedirect.com/science/article/pii/S1462901118305768> (Accessed March 29,
29 2019).
- 30 Singh, B. P., and A. L. Cowie, 2014: Long-term influence of biochar on native organic carbon
31 mineralisation in a low-carbon clayey soil. *Sci. Rep.*, **4**, 3687,
32 doi:10.1038/srep03687 <https://www.nature.com/articles/srep03687#supplementary-information>.
33 <http://dx.doi.org/10.1038/srep03687>.
- 34 Singh, B. P., B. J. Hatton, S. Balwant, and a L. Cowie, 2010: The role of biochar in reducing nitrous
35 oxide emissions and nitrogen leaching from soil. *Proc. 19th World Congr. Soil Sci. Soil Solut. a*
36 *Chang. world, Brisbane, Aust. 1-6 August 2010. Congr. Symp. 4 Greenh. gases from soils*, 257–
37 259, doi:10.2134/jeq2009.0138.
- 38 Singh, B. P., A. L. Cowie, and R. J. Smernik, 2012: Biochar Carbon Stability in a Clayey Soil As a
39 Function of Feedstock and Pyrolysis Temperature. *Environ. Sci. Technol.*,
40 doi:10.1021/es302545b. <http://pubs.acs.org/doi/abs/10.1021/es302545b>.
- 41 —, Y. Fang, M. Boersma, D. Collins, L. Van Zwieten, and L. M. Macdonald, 2015: In Situ
42 Persistence and Migration of Biochar Carbon and Its Impact on Native Carbon Emission in
43 Contrasting Soils under Managed Temperate Pastures. *PLoS One*, **10**, e0141560,
44 doi:10.1371/journal.pone.0141560. <http://dx.plos.org/10.1371/journal.pone.0141560> (Accessed
45 December 1, 2017).
- 46 Sinha, D., and M. R. Ray, 2015: Health Effects of Indoor Air Pollution Due to Cooking with Biomass
47 Fuel. Humana Press, Cham, 267–302 http://link.springer.com/10.1007/978-3-319-19096-9_14
48 (Accessed April 10, 2019).
- 49 Siry, J. P., F. W. Cabbage, and M. R. Ahmed, 2005: Sustainable forest management: Global trends

- 1 and opportunities. *For. Policy Econ.*, **7**, 551–561, doi:10.1016/j.forpol.2003.09.003.
- 2 Sklenicka, P., K. J. Molnarova, M. Salek, P. Simova, J. Vlasak, P. Sekac, and V. Janovska, 2015:
3 Owner or tenant: Who adopts better soil conservation practices? *Land use policy*, **47**, 253–261,
4 doi:10.1016/J.LANDUSEPOL.2015.04.017.
5 <https://www.sciencedirect.com/science/article/pii/S0264837715001192> (Accessed April 3,
6 2019).
- 7 Skoufias, E., M. Rabassa, and S. Olivieri, 2011: *The poverty impacts of climate change: A review of*
8 *the evidence*. The World Bank, <http://elibrary.worldbank.org/doi/book/10.1596/1813-9450-5622>
9 (Accessed March 8, 2019).
- 10 Ben Slimane, A., D. Raclot, O. Evrard, M. Sanaa, I. Lefevre, and Y. Le Bissonnais, 2016: Relative
11 Contribution of Rill/Interrill and Gully/Channel Erosion to Small Reservoir Siltation in
12 Mediterranean Environments. *L. Degrad. Dev.*, **27**, 785–797, doi:10.1002/ldr.2387.
13 <http://doi.wiley.com/10.1002/ldr.2387> (Accessed March 8, 2019).
- 14 Sloan, S., and J. A. Sayer, 2015: Forest Resources Assessment of 2015 shows positive global trends
15 but forest loss and degradation persist in poor tropical countries. *For. Ecol. Manage.*, **352**, 134–
16 145, doi:10.1016/J.FORECO.2015.06.013.
17 <https://www.sciencedirect.com/science/article/pii/S0378112715003394> (Accessed October 31,
18 2018).
- 19 Slobbe, E., H. J. Vriend, S. Aarninkhof, K. Lulofs, M. Vries, and P. Dircke, 2013: Building with
20 Nature: in search of resilient storm surge protection strategies. *Nat. Hazards*, **65**, 947–966,
21 doi:10.1007/s11069-012-0342-y.
- 22 Smeets, E., F. X. Johnson, and G. Ballard-Tremere, 2012: Traditional and Improved Use of Biomass
23 for Energy in Africa. *Bioenergy for Sustainable Development in Africa*, R. Janssen and D. Rutz,
24 Eds., Springer Netherlands, Dordrecht, 3–12.
- 25 Smith, H. E., F. Eigenbrod, D. Kafumbata, M. D. Hudson, and K. Schreckenberger, 2015: Criminals by
26 necessity: the risky life of charcoal transporters in Malawi. *For. Trees Livelihoods*, **24**, 259–274,
27 doi:10.1080/14728028.2015.1062808.
28 <http://www.tandfonline.com/doi/full/10.1080/14728028.2015.1062808> (Accessed November 2,
29 2018).
- 30 Smith, J., D. R. Nayak, and P. Smith, 2012: Renewable energy: Avoid constructing wind farms on
31 peat. *Nature*, doi:10.1038/489033d.
- 32 Smith, P., 2016: Soil carbon sequestration and biochar as negative emission technologies. *Glob.*
33 *Chang. Biol.*, **22**, 1315–1324, doi:10.1111/gcb.13178. <http://doi.wiley.com/10.1111/gcb.13178>
34 (Accessed December 1, 2017).
- 35 —, and Coauthors, 2016a: Global change pressures on soils from land use and management. *Glob.*
36 *Chang. Biol.*, **22**, 1008–1028, doi:10.1111/gcb.13068. <http://doi.wiley.com/10.1111/gcb.13068>
37 (Accessed May 21, 2018).
- 38 —, and Coauthors, 2016b: Biophysical and economic limits to negative CO₂ emissions. *Nat. Clim.*
39 *Chang.*, **6**, 42–50, doi:10.1038/nclimate2870. <http://www.nature.com/articles/nclimate2870>
40 (Accessed March 16, 2018).
- 41 Smith, P. and M. B., and Coauthors, 2014: Agriculture, Forestry and Other Land Use (AFOLU).
42 *Climate Change 2014: Mitigation of Climate Change. Contribution of Working Group III to the*
43 *Fifth Assessment Report of the Intergovernmental Panel on Climate Change*, O. Edenhofer et
44 al., Eds., Cambridge University Press, Cambridge, United Kingdom and New York, NY, USA,
45 p. 83.
- 46 Smyth, C., W. A. Kurz, G. Rampley, T. C. Lemprière, and O. Schwab, 2017: Climate change
47 mitigation potential of local use of harvest residues for bioenergy in Canada. *GCB Bioenergy*, **9**,
48 817–832, doi:10.1111/gcbb.12387.

- 1 Smyth, C. E. E., G. Stinson, E. Neilson, T. C. C. Lemprière, M. Hafer, G. J. J. Rampley, and W. A. A.
2 Kurz, 2014: Quantifying the biophysical climate change mitigation potential of Canada's forest
3 sector. *Biogeosciences*, **11**, 3515–3529, doi:10.5194/bg-11-3515-2014.
4 www.biogeosciences.net/11/3515/2014/ (Accessed May 26, 2018).
- 5 Sobel, A. H., S. J. Camargo, T. M. Hall, C.-Y. Lee, M. K. Tippet, and A. A. Wing, 2016: Human
6 influence on tropical cyclone intensity. *Science*, **353**, 242–246, doi:10.1126/science.aaf6574.
7 <http://www.ncbi.nlm.nih.gov/pubmed/27418502> (Accessed April 10, 2019).
- 8 Sokolik, I. N., and O. B. Toon, 1996: Direct radiative forcing by anthropogenic airborne mineral
9 aerosols. *Nature*, **381**, 681–683, doi:10.1038/381681a0.
10 <http://www.nature.com/doi/10.1038/381681a0> (Accessed October 3, 2018).
- 11 Solly, E. F., B. D. Lindahl, M. A. Dawes, M. Peter, R. C. Souza, C. Rixen, and F. Hagedorn, 2017:
12 Experimental soil warming shifts the fungal community composition at the alpine treeline. *New*
13 *Phytol.*, **215**, 766–778, doi:10.1111/nph.14603. <http://doi.wiley.com/10.1111/nph.14603>
14 (Accessed March 8, 2019).
- 15 Solomon, N., E. Birhane, C. Gordon, M. Haile, F. Taheri, H. Azadi, and J. Scheffran, 2018:
16 Environmental impacts and causes of conflict in the Horn of Africa: A review. *Earth-Science*
17 *Rev.*, **177**, 284–290, doi:10.1016/J.EARSCIREV.2017.11.016.
18 <https://www.sciencedirect.com/science/article/pii/S0012825217301356> (Accessed May 16,
19 2018).
- 20 Song, X.-P., M. C. Hansen, S. V. Stehman, P. V. Potapov, A. Tyukavina, E. F. Vermote, and J. R.
21 Townshend, 2018: Global land change from 1982 to 2016. *Nature*, **560**, 639–643,
22 doi:10.1038/s41586-018-0411-9. <http://www.nature.com/articles/s41586-018-0411-9> (Accessed
23 October 29, 2018).
- 24 Sonneveld, B. G. J. S., and D. L. Dent, 2009: How good is GLASOD? *J. Environ. Manage.*, **90**, 274–
25 283, doi:10.1016/J.JENVMAN.2007.09.008.
26 <https://www.sciencedirect.com/science/article/pii/S0301479707003441> (Accessed May 11,
27 2018).
- 28 Sonwa, D. J., S. Walker, R. Nasi, and M. Kanninen, 2011a: Potential synergies of the main current
29 forestry efforts and climate change mitigation in Central Africa. *Sustain. Sci.*, **6**, 59–67,
30 doi:10.1007/s11625-010-0119-8. <http://link.springer.com/10.1007/s11625-010-0119-8>
31 (Accessed March 22, 2019).
- 32 —, —, —, and —, 2011b: Potential synergies of the main current forestry efforts and
33 climate change mitigation in Central Africa. *Sustain. Sci.*, **6**, 59–67, doi:10.1007/s11625-010-
34 0119-8. <http://link.springer.com/10.1007/s11625-010-0119-8> (Accessed March 26, 2019).
- 35 —, S. F. Weise, G. Schroth, M. J. J. Janssens, and Howard-Yana Shapiro, 2014: Plant diversity
36 management in cocoa agroforestry systems in West and Central Africa—effects of markets and
37 household needs. *Agrofor. Syst.*, **88**, 1021–1034, doi:10.1007/s10457-014-9714-5.
- 38 —, —, B. A. Nkongmeneck, M. Tchatat, and M. J. J. Janssens, 2017: Structure and composition
39 of cocoa agroforests in the humid forest zone of Southern Cameroon. *Agrofor. Syst.*, **91**, 451–
40 470, doi:10.1007/s10457-016-9942-y.
- 41 Sonwa, F., S. F. Weise, M. Tchatat, B. A. Nkongmeneck, A. A. Adesina, O. Ndoeye, and J.
42 Gockowski, 2001: *The role of cocoa agroforests in rural and community forestry in Southern*
43 *Cameroon*. Overseas Development Institute, London, 1-10 pp.
44 <https://www.odi.org/resources/docs/1226.pdf>.
- 45 Soule, M. J., A. Tegene, and K. D. Wiebe, 2000: Land Tenure and the Adoption of Conservation
46 Practices. *Am. J. Agric. Econ.*, **82**, 993–1005, doi:10.1111/0002-9092.00097.
47 <https://academic.oup.com/ajae/article-lookup/doi/10.1111/0002-9092.00097> (Accessed February
48 28, 2019).
- 49 Sousa, F. F. de, C. Vieira-da-Silva, and F. B. Barros, 2018: The (in)visible market of miriti (Mauritia

- 1 flexuosa L.f.) fruits, the “winter acai”, in Amazonian riverine communities of Abaetetuba,
2 Northern Brazil. *Glob. Ecol. Conserv.*, **14**, e00393, doi:10.1016/j.gecco.2018.e00393.
- 3 Soussana, J.-F., and G. Lemaire, 2014: Coupling carbon and nitrogen cycles for environmentally
4 sustainable intensification of grasslands and crop-livestock systems. *Agric. Ecosyst. Environ.*,
5 **190**, 9–17, doi:10.1016/J.AGEE.2013.10.012.
6 <https://www.sciencedirect.com/science/article/pii/S0167880913003526> (Accessed October 18,
7 2018).
- 8 Soussana, J.-F., P. Loiseau, N. Vuichard, E. Ceschia, J. Balesdent, T. Chevallier, and D. Arrouays,
9 2006: Carbon cycling and sequestration opportunities in temperate grasslands. *Soil Use Manag.*,
10 **20**, 219–230, doi:10.1111/j.1475-2743.2004.tb00362.x. [http://doi.wiley.com/10.1111/j.1475-](http://doi.wiley.com/10.1111/j.1475-2743.2004.tb00362.x)
11 [2743.2004.tb00362.x](http://doi.wiley.com/10.1111/j.1475-2743.2004.tb00362.x) (Accessed October 18, 2018).
- 12 Soussana, J.-F., and Coauthors, 2019: Matching policy and science: Rationale for the ‘4 per 1000 -
13 soils for food security and climate’ initiative. *Soil Tillage Res.*, **188**, 3–15,
14 doi:10.1016/J.STILL.2017.12.002.
15 <https://www.sciencedirect.com/science/article/pii/S0167198717302271> (Accessed April 13,
16 2019).
- 17 Sovacool, B. K., 2012: Perceptions of climate change risks and resilient island planning in the
18 Maldives. *Mitig. Adapt. Strateg. Glob. Chang.*, **17**, 731–752, doi:10.1007/s11027-011-9341-7.
- 19 Sparrevik, M., C. Adam, V. Martinsen, Jubaedah, and G. Cornelissen, 2015: Emissions of gases and
20 particles from charcoal/biochar production in rural areas using medium-sized traditional and
21 improved “retort” kilns. *Biomass and Bioenergy*, **72**, 65–73,
22 doi:<https://doi.org/10.1016/j.biombioe.2014.11.016>.
23 <http://www.sciencedirect.com/science/article/pii/S0961953414005170>.
- 24 Specht, M. J., S. R. R. Pinto, U. P. Albuquerque, M. Tabarelli, and F. P. L. Melo, 2015: Burning
25 biodiversity: Fuelwood harvesting causes forest degradation in human-dominated tropical
26 landscapes. *Glob. Ecol. Conserv.*, **3**, 200–209, doi:<https://doi.org/10.1016/j.gecco.2014.12.002>.
27 <http://www.sciencedirect.com/science/article/pii/S2351989414000894>.
- 28 Spence, J. R., 2001: The new boreal forestry: adjusting timber management to accommodate
29 biodiversity. *Trends Ecol. Evol.*, **16**, 591–593, doi:10.1016/S0169-5347(01)02335-7.
30 <https://www.sciencedirect.com/science/article/pii/S0169534701023357> (Accessed May 26,
31 2018).
- 32 Spencer, T., and Coauthors, 2016: Global coastal wetland change under sea-level rise and related
33 stresses: The DIVA Wetland Change Model. *Glob. Planet. Change*, **139**, 15–30,
34 doi:10.1016/J.GLOPLACHA.2015.12.018.
35 <https://www.sciencedirect.com/science/article/pii/S0921818115301879> (Accessed March 14,
36 2019).
- 37 Sprunger, C. D., S. W. Culman, G. P. Robertson, and S. S. Snapp, 2018: Perennial grain on a Midwest
38 Alfisol shows no sign of early soil carbon gain. *Renew. Agric. Food Syst.*, **33**, 360–372,
39 doi:10.1017/S1742170517000138.
40 https://www.cambridge.org/core/product/identifier/S1742170517000138/type/journal_article
41 (Accessed April 3, 2019).
- 42 St.Clair, S. B., and J. P. Lynch, 2010: The opening of Pandora’s Box: climate change impacts on soil
43 fertility and crop nutrition in developing countries. *Plant Soil*, **335**, 101–115,
44 doi:10.1007/s11104-010-0328-z. <http://link.springer.com/10.1007/s11104-010-0328-z>
45 (Accessed May 15, 2018).
- 46 Stafford Smith, D. M., G. M. McKeon, I. W. Watson, B. K. Henry, G. S. Stone, W. B. Hall, and S. M.
47 Howden, 2007: Learning from episodes of degradation and recovery in variable Australian
48 rangelands. *Proc. Natl. Acad. Sci. U. S. A.*, **104**, 20690–20695, doi:10.1073/pnas.0704837104.
49 <http://www.ncbi.nlm.nih.gov/pubmed/18093932> (Accessed March 1, 2019).

- 1 ter Steege, H., and Coauthors, 2013: Hyperdominance in the Amazonian Tree Flora. *Science* (80-.),
2 **342**, 1243092–1243092, doi:10.1126/science.1243092.
- 3 Steffen, W. L., and Coauthors, 2005: *Global change and the earth system : a planet under pressure*.
4 Springer, Berlin, Germany, 336 pp.
- 5 Steinkamp, J., and T. Hickler, 2015: Is drought-induced forest dieback globally increasing? *J. Ecol.*,
6 **103**, 31–43, doi:10.1111/1365-2745.12335. <http://doi.wiley.com/10.1111/1365-2745.12335>
7 (Accessed March 14, 2019).
- 8 Stevens, P., T. Roberts, and S. Lucas, 2015: Life on mars: Using micro-topographic relief to secure
9 soil, water and biocapacity. Barton ACT, Australia, Engineers Australia
10 <https://search.informit.com.au/documentSummary;dn=732466230109912;res=IELENG>
11 (Accessed April 14, 2019).
- 12 Steward, P. R., A. J. Dougill, C. Thierfelder, C. M. Pittelkow, L. C. Stringer, M. Kudzala, and G. E.
13 Shackelford, 2018: The adaptive capacity of maize-based conservation agriculture systems to
14 climate stress in tropical and subtropical environments: A meta-regression of yields. *Agric.*
15 *Ecosyst. Environ.*, **251**, 194–202, doi:10.1016/J.AGEE.2017.09.019.
16 <https://www.sciencedirect.com/science/article/pii/S016788091730419X> (Accessed May 14,
17 2018).
- 18 Stocking, M. A., N. Murnaghan, and N. Murnaghan, 2001: *A Handbook for the Field Assessment of*
19 *Land Degradation*. Routledge, <https://www.taylorfrancis.com/books/9781849776219> (Accessed
20 April 28, 2018).
- 21 Stoorvogel, J. J., M. Bakkenes, A. J. A. M. Temme, N. H. Batjes, and B. J. E. ten Brink, 2017: S-
22 World: A Global Soil Map for Environmental Modelling. *L. Degrad. Dev.*, **28**, 22–33,
23 doi:10.1002/ldr.2656. <http://doi.wiley.com/10.1002/ldr.2656> (Accessed April 12, 2019).
- 24 Strefler, J., T. Amann, N. Bauer, E. Krieglner, and J. Hartmann, 2018: Potential and costs of carbon
25 dioxide removal by enhanced weathering of rocks. *Environ. Res. Lett.*, **13**, 034010,
26 doi:10.1088/1748-9326/aaa9c4. [http://stacks.iop.org/1748-
27 9326/13/i=3/a=034010?key=crossref.9d7613c7d9e3b6224a40aa812b709bfe](http://stacks.iop.org/1748-9326/13/i=3/a=034010?key=crossref.9d7613c7d9e3b6224a40aa812b709bfe) (Accessed April 12,
28 2019).
- 29 Stringer, L. C., M. Akhtar-Schuster, M. J. Marques, F. Amiraslani, S. Quatrini, and E. M. Abraham,
30 2011: Combating land degradation and desertification and enhancing food security: Towards
31 integrated solutions. *Ann. Arid Zone*, **50**, 1–23.
32 <https://www.researchgate.net/publication/259459727> (Accessed May 15, 2018).
- 33 Strunz, S., 2012: Is conceptual vagueness an asset? Arguments from philosophy of science applied to
34 the concept of resilience. *Ecol. Econ.*, **76**, 112–118, doi:10.1016/J.ECOLECON.2012.02.012.
35 <https://www.sciencedirect.com/science/article/pii/S0921800912000766> (Accessed March 30,
36 2019).
- 37 Sturm, M., 2005: Changing snow and shrub conditions affect albedo with global implications. *J.*
38 *Geophys. Res.*, **110**, G01004, doi:10.1029/2005JG000013.
39 <https://agupubs.onlinelibrary.wiley.com/doi/abs/10.1029/2005JG000013> (Accessed October 3,
40 2018).
- 41 Sturrock, R. N., S. J. Frankel, A. V. Brown, P. E. Hennon, J. T. Kliejunas, K. J. Lewis, J. J. Worrall,
42 and A. J. Woods, 2011: Climate change and forest diseases. *Plant Pathol.*, **60**, 133–149,
43 doi:10.1111/j.1365-3059.2010.02406.x. <http://doi.wiley.com/10.1111/j.1365-3059.2010.02406.x>
44 (Accessed March 5, 2018).
- 45 Sufo Kankeu, R., D. J. Sonwa, R. Eba'a Atyi, and N. M. Moankang Nkal, 2016: Quantifying post
46 logging biomass loss using satellite images and ground measurements in Southeast Cameroon. *J.*
47 *For. Res.*, **27**, 1415–1426, doi:10.1007/s11676-016-0277-3.
48 <http://link.springer.com/10.1007/s11676-016-0277-3> (Accessed October 31, 2018).
- 49 Sulaiman, C., A. S. Abdul-Rahim, H. O. Mohd-Shahwahid, and L. Chin, 2017: Wood fuel

- 1 consumption, institutional quality, and forest degradation in sub-Saharan Africa: Evidence from
2 a dynamic panel framework. *Ecol. Indic.*, **74**, 414–419,
3 doi:<https://doi.org/10.1016/j.ecolind.2016.11.045>.
4 <http://www.sciencedirect.com/science/article/pii/S1470160X16306938>.
- 5 Sundström, J. F., and Coauthors, 2014: Future threats to agricultural food production posed by
6 environmental degradation, climate change, and animal and plant diseases – a risk analysis in
7 three economic and climate settings. *Food Secur.*, **6**, 201–215, doi:10.1007/s12571-014-0331-y.
8 <http://link.springer.com/10.1007/s12571-014-0331-y> (Accessed May 14, 2018).
- 9 Sussams, L. W., W. R. Sheate, and R. P. Eales, 2015: Green infrastructure as a climate change
10 adaptation policy intervention: Muddying the waters or clearing a path to a more secure future?
11 *J. Environ. Manage.*, **147**, 184–193, doi:10.1016/J.JENVMAN.2014.09.003.
12 <https://www.sciencedirect.com/science/article/pii/S0301479714004411> (Accessed May 17,
13 2018).
- 14 Swails, E., D. Jaye, L. Verchot, K. Hergoualc’h, M. Schirrmann, N. Borchard, N. Wahyuni, and D.
15 Lawrence, 2018: Will CO2 Emissions from Drained Tropical Peatlands Decline Over Time?
16 Links Between Soil Organic Matter Quality, Nutrients, and C Mineralization Rates. *Ecosystems*,
17 **21**, 868–885, doi:10.1007/s10021-017-0190-4.
- 18 Sydeman, W. J., M. García-Reyes, D. S. Schoeman, R. R. Rykaczewski, S. A. Thompson, B. A.
19 Black, and S. J. Bograd, 2014: Climate change. Climate change and wind intensification in
20 coastal upwelling ecosystems. *Science*, **345**, 77–80, doi:10.1126/science.1251635.
21 <http://www.ncbi.nlm.nih.gov/pubmed/24994651> (Accessed October 24, 2018).
- 22 Szabo, S., and Coauthors, 2016: Population dynamics, delta vulnerability and environmental change:
23 comparison of the Mekong, Ganges–Brahmaputra and Amazon delta regions. *Sustain. Sci.*, **11**,
24 539–554, doi:10.1007/s11625-016-0372-6. <http://link.springer.com/10.1007/s11625-016-0372-6>
25 (Accessed March 14, 2019).
- 26 Tacoli, C., 2009: Crisis or adaptation? Migration and climate change in a context of high mobility.
27 *Environ. Urban.*, **21**, 513–525, doi:10.1177/0956247809342182.
28 <http://journals.sagepub.com/doi/10.1177/0956247809342182> (Accessed November 2, 2018).
- 29 Tadesse, G., 2001: Land Degradation: A Challenge to Ethiopia. *Environ. Manage.*, **27**, 815–824,
30 doi:10.1007/s002670010190.
31 <https://link.springer.com/content/pdf/10.1007%2Fs002670010190.pdf> (Accessed March 15,
32 2018).
- 33 Tadesse, G., B. Algieri, M. Kalkuhl, and J. von Braun, 2014: Drivers and triggers of international
34 food price spikes and volatility. *Food Policy*, **47**, 117–128,
35 doi:10.1016/J.FOODPOL.2013.08.014.
36 <https://www.sciencedirect.com/science/article/pii/S0306919213001188> (Accessed March 11,
37 2019).
- 38 Tamarin-Brodsky, T., and Y. Kaspi, 2017: Enhanced poleward propagation of storms under climate
39 change. *Nat. Geosci.*, **10**, 908–913, doi:10.1038/s41561-017-0001-8.
40 <http://www.nature.com/articles/s41561-017-0001-8> (Accessed March 16, 2018).
- 41 Tang, J., S. Luysaert, A. D. Richardson, W. Kutsch, and I. A. Janssens, 2014: Steeper declines in
42 forest photosynthesis than respiration explain age-driven decreases in forest growth. *Proc. Natl.*
43 *Acad. Sci.*, **111**, 8856–8860, doi:10.1073/pnas.1320761111.
- 44 Tarfasa, S., B. B. Balana, T. Tefera, T. Woldeamanuel, A. Moges, M. Dinato, and H. Black, 2018:
45 Modeling Smallholder Farmers’ Preferences for Soil Management Measures: A Case Study
46 From South Ethiopia. *Ecol. Econ.*, **145**, 410–419, doi:10.1016/j.ecolecon.2017.11.027.
47 <http://linkinghub.elsevier.com/retrieve/pii/S092180091631583X> (Accessed April 16, 2018).
- 48 Van Tassel, D. L., and Coauthors, 2017: Accelerating silphium domestication: An opportunity to
49 develop new crop ideotypes and breeding strategies informed by multiple disciplines. *Crop Sci.*,

- 1 **57**, 1274–1284, doi:10.2135/cropsci2016.10.0834.
- 2 Taufik, M., P. J. J. F. Torfs, R. Uijlenhoet, P. D. Jones, D. Murdiyarso, and H. A. J. Van Lanen, 2017:
3 Amplification of wildfire area burnt by hydrological drought in the humid tropics. *Nat. Clim.*
4 *Chang.*, **7**, 428–431, doi:10.1038/nclimate3280. <http://www.nature.com/articles/nclimate3280>
5 (Accessed March 16, 2019).
- 6 Taylor, A. R., M. Seedre, B. W. Brassard, and H. Y. H. Chen, 2014: Decline in Net Ecosystem
7 Productivity Following Canopy Transition to Late-Succession Forests. *Ecosystems*, **17**, 778–
8 791, doi:10.1007/s10021-014-9759-3.
- 9 Taylor, L. L., D. J. Beerling, S. Quegan, and S. A. Banwart, 2017: Simulating carbon capture by
10 enhanced weathering with croplands: an overview of key processes highlighting areas of future
11 model development. *Biol. Lett.*, **13**, 20160868, doi:10.1098/rsbl.2016.0868.
12 <http://rsbl.royalsocietypublishing.org/lookup/doi/10.1098/rsbl.2016.0868> (Accessed April 12,
13 2019).
- 14 Tengberg, A., and S. Valencia, 2018: Integrated approaches to natural resources management-Theory
15 and practice. *L. Degrad. Dev.*, **29**, 1845–1857, doi:10.1002/ldr.2946.
16 <http://doi.wiley.com/10.1002/ldr.2946> (Accessed October 5, 2018).
- 17 Tengberg, A., S. Fredholm, I. Eliasson, I. K.-E. Services, and U. 2012, 2012: Cultural ecosystem
18 services provided by landscapes: Assessment of heritage values and identity. *Ecosyst. Serv.*, **2**,
19 14–26. <https://www.sciencedirect.com/science/article/pii/S2212041612000113> (Accessed
20 February 28, 2019).
- 21 Tengberg, A., F. Radstake, K. Zhang, and B. Dunn, 2016: Scaling up of Sustainable Land
22 Management in the Western People’s Republic of China: Evaluation of a 10-Year Partnership.
23 *L. Degrad. Dev.*, doi:10.1002/ldr.2270.
- 24 Ter-Mikaelian, M. T., S. J. Colombo, and J. Chen, 2013: Effects of harvesting on spatial and temporal
25 diversity of carbon stocks in a boreal forest landscape. *Ecol. Evol.*, **3**, 3738–3750,
26 doi:10.1002/ece3.751.
- 27 —, —, and —, 2014: The Burning Question: Does Forest Bioenergy Reduce Carbon
28 Emissions? A Review of Common Misconceptions about Forest Carbon Accounting. *J. For.*,
29 **113**, 57–68, doi:10.5849/jof.14-016.
- 30 Terrer, C., S. Vicca, B. A. Hungate, R. P. Phillips, and I. C. Prentice, 2016: Mycorrhizal association
31 as a primary control of the CO₂ fertilization effect. *Science*, **353**, 72–74,
32 doi:10.1126/science.aaf4610. <http://www.ncbi.nlm.nih.gov/pubmed/27365447> (Accessed March
33 4, 2019).
- 34 Terry, J. P., and A. Y. A. Lau, 2018: Magnitudes of nearshore waves generated by tropical cyclone
35 Winston, the strongest landfalling cyclone in South Pacific records. Unprecedented or
36 unremarkable? *Sediment. Geol.*, **364**, 276–285, doi:10.1016/J.SEDGEO.2017.10.009.
37 <https://www.sciencedirect.com/science/article/pii/S0037073817302282> (Accessed April 11,
38 2019).
- 39 Tesfaye, A., R. Brouwer, P. van der Zaag, and W. Negatu, 2016: Assessing the costs and benefits of
40 improved land management practices in three watershed areas in Ethiopia. *Int. Soil Water*
41 *Conserv. Res.*, **4**, 20–29, doi:10.1016/J.ISWCR.2016.01.003.
42 <https://www.sciencedirect.com/science/article/pii/S2095633915301015> (Accessed April 17,
43 2019).
- 44 Teshome, A., J. de Graaff, C. Ritsema, and M. Kassie, 2016: Farmers’ Perceptions about the
45 Influence of Land Quality, Land Fragmentation and Tenure Systems on Sustainable Land
46 Management in the North Western Ethiopian Highlands. *L. Degrad. Dev.*, **27**, 884–898,
47 doi:10.1002/ldr.2298. <http://doi.wiley.com/10.1002/ldr.2298> (Accessed April 17, 2019).
- 48 Tessler, Z. D., C. J. Vörösmarty, M. Grossberg, I. Gladkova, H. Aizenman, J. P. M. Syvitski, and E.

- 1 Foufoula-Georgiou, 2015: ENVIRONMENTAL SCIENCE. Profiling risk and sustainability in
2 coastal deltas of the world. *Science*, **349**, 638–643, doi:10.1126/science.aab3574.
3 <http://www.ncbi.nlm.nih.gov/pubmed/26250684> (Accessed October 27, 2018).
- 4 Tessler, Z. D., C. J. Vörösmarty, M. Grossberg, I. Gladkova, and H. Aizenman, 2016: A global
5 empirical typology of anthropogenic drivers of environmental change in deltas. *Sustain. Sci.*, **11**,
6 525–537, doi:10.1007/s11625-016-0357-5. <http://link.springer.com/10.1007/s11625-016-0357-5>
7 (Accessed October 27, 2018).
- 8 Testa, S., K. Soudani, L. Boschetti, and E. Borgogno Mondino, 2018: MODIS-derived EVI, NDVI
9 and WDRVI time series to estimate phenological metrics in French deciduous forests. *Int. J.*
10 *Appl. Earth Obs. Geoinf.*, **64**, 132–144, doi:10.1016/j.jag.2017.08.006.
11 <https://www.sciencedirect.com/science/article/pii/S030324341730171X> (Accessed March 18,
12 2019).
- 13 Thies, J. E., M. C. Rillig, and E. R. Graber, 2015: Biochar effects on the abundance, activity and
14 diversity of the soil biota. *Biochar for Environmental Management: Science, Technology and*
15 *Implementation*.
- 16 Thompson, I., B. Mackey, S. McNulty, and A. Mosseler, 2009: Forest resilience, biodiversity, and
17 climate change. *A synthesis of the biodiversity/resilience/stability relationship in forest*
18 *ecosystems. Secretariat of the Convention on Biological Diversity, Montreal. Technical Series,*
19 *Vol. 43 of, 67.*
- 20 Thomson, J., and W. E. Rogers, 2014: Swell and sea in the emerging Arctic Ocean. *Geophys. Res.*
21 *Let.*, **41**, 3136–3140, doi:10.1002/2014GL059983.
22 <http://doi.wiley.com/10.1002/2014GL059983> (Accessed March 14, 2019).
- 23 Thomson, L. J., S. Macfadyen, and A. A. Hoffmann, 2010: Predicting the effects of climate change on
24 natural enemies of agricultural pests. *Biol. Control*, **52**, 296–306,
25 doi:10.1016/J.BIOCONTROL.2009.01.022.
26 <https://www.sciencedirect.com/science/article/pii/S1049964409000413> (Accessed May 14,
27 2018).
- 28 Thorén, H., and L. Olsson, 2018: Is resilience a normative concept? *Resilience*, **6**, 112–128,
29 doi:10.1080/21693293.2017.1406842.
30 <https://www.tandfonline.com/doi/full/10.1080/21693293.2017.1406842> (Accessed May 25,
31 2018).
- 32 Tian, F., M. Brandt, Y. Y. Liu, K. Rasmussen, and R. Fensholt, 2017: Mapping gains and losses in
33 woody vegetation across global tropical drylands. *Glob. Chang. Biol.*, **23**, 1748–1760,
34 doi:10.1111/gcb.13464. <http://doi.wiley.com/10.1111/gcb.13464> (Accessed December 20,
35 2017).
- 36 Tian, H., and Coauthors, 2015: North American terrestrial CO₂ uptake largely offset by CH₄ and
37 N₂O emissions: toward a full accounting of the greenhouse gas budget. *Clim. Change*, **129**,
38 413–426, doi:10.1007/s10584-014-1072-9. <http://link.springer.com/10.1007/s10584-014-1072-9>
39 (Accessed November 2, 2018).
- 40 Tighe, M., C. Muñoz-Robles, N. Reid, B. Wilson, and S. V Briggs, 2012: Hydrological thresholds of
41 soil surface properties identified using conditional inference tree analysis. *Earth Surf. Process.*
42 *Landforms*, **37**, 620–632, doi:10.1002/esp.3191. <http://dx.doi.org/10.1002/esp.3191>.
- 43 Tilman, D., C. Balzer, J. Hill, and B. L. Befort, 2011: Global food demand and the sustainable
44 intensification of agriculture. *Proc. Natl. Acad. Sci. U. S. A.*, **108**, 20260–20264,
45 doi:10.1073/pnas.1116437108. <http://www.ncbi.nlm.nih.gov/pubmed/22106295> (Accessed
46 March 15, 2018).
- 47 Tonosaki K, Kawai S, T. K., 2014: Cooling Potential of Urban Green Spaces in Summer. *Designing*
48 *Low Carbon Societies in Landscapes. Editors Nakagoshi N, Mabuhay AJ.*, Springer, Tokyo, 15–
49 34.

- 1 Torquebiau, E., 1992: Are tropical agroforestry home gardens sustainable? *Agric. Ecosyst. Environ.*,
2 **41**, 189–207, doi:10.1016/0167-8809(92)90109-O.
3 <https://www.sciencedirect.com/science/article/pii/016788099290109O> (Accessed October 29,
4 2018).
- 5 Toth, G. G., P. K. Ramachandran Nair, M. Jacobson, Y. Widyaningsih, and C. P. Duffy, 2017:
6 Malawi's Energy Needs and Agroforestry: Adoption Potential of Woodlots. *Hum. Ecol.*, **45**,
7 735–746, doi:10.1007/s10745-017-9944-z. <http://link.springer.com/10.1007/s10745-017-9944-z>
8 (Accessed October 22, 2018).
- 9 Toulmin, C., 2009: Securing land and property rights in sub-Saharan Africa: The role of local
10 institutions. *Land use policy*, **26**, 10–19, doi:10.1016/J.LANDUSEPOL.2008.07.006.
11 <https://www.sciencedirect.com/science/article/pii/S0264837708000811> (Accessed April 16,
12 2018).
- 13 —, and K. Brock, 2016: Desertification in the Sahel: Local Practice Meets Global Narrative.
14 Springer, Berlin, Heidelberg, 37–63 http://link.springer.com/10.1007/978-3-642-16014-1_2
15 (Accessed May 12, 2018).
- 16 Trahan, M. W., and B. A. Schubert, 2016: Temperature-induced water stress in high-latitude forests in
17 response to natural and anthropogenic warming. *Glob. Chang. Biol.*, **22**, 782–791,
18 doi:10.1111/gcb.13121. <http://doi.wiley.com/10.1111/gcb.13121> (Accessed March 14, 2019).
- 19 Trenberth, K. E., 1999: Conceptual Framework for Changes of Extremes of the Hydrological Cycle
20 With Climate Change. *Weather and Climate Extremes*, Springer Netherlands, Dordrecht, 327–
21 339 http://link.springer.com/10.1007/978-94-015-9265-9_18 (Accessed November 14, 2017).
- 22 —, 2011: Changes in precipitation with climate change. *Clim. Res.*, **47**, 123–138,
23 doi:10.2307/24872346. <http://www.jstor.org/stable/24872346> (Accessed March 9, 2018).
- 24 Trofymow, J. A., G. Stinson, and W. A. Kurz, 2008: Derivation of a spatially explicit 86-year
25 retrospective carbon budget for a landscape undergoing conversion from old-growth to managed
26 forests on Vancouver Island, BC. *For. Ecol. Manage.*, **256**, doi:10.1016/j.foreco.2008.02.056.
- 27 Trumbore, S., P. Brando, and H. Hartmann, 2015: Forest health and global change. *Science*, **349**, 814–
28 818, doi:10.1126/science.aac6759. <http://www.ncbi.nlm.nih.gov/pubmed/26293952> (Accessed
29 October 27, 2018).
- 30 Tu, S., F. Xu, and J. Xu, 2018: Regime shift in the destructiveness of tropical cyclones over the
31 western North Pacific. *Environ. Res. Lett.*, **13**, 094021, doi:10.1088/1748-9326/aade3a.
32 [http://stacks.iop.org/1748-
33 9326/13/i=9/a=094021?key=crossref.784ca4128f17cd025d409d28964630a8](http://stacks.iop.org/1748-9326/13/i=9/a=094021?key=crossref.784ca4128f17cd025d409d28964630a8) (Accessed October
34 14, 2018).
- 35 Turetsky, M. R., and Coauthors, 2014: A synthesis of methane emissions from 71 northern, temperate,
36 and subtropical wetlands. *Glob. Chang. Biol.*, **20**, 2183–2197, doi:10.1111/gcb.12580.
37 <http://doi.wiley.com/10.1111/gcb.12580> (Accessed October 3, 2018).
- 38 Turner, B. L., and J. A. Sabloff, 2012: Classic Period collapse of the Central Maya Lowlands: insights
39 about human-environment relationships for sustainability. *Proc. Natl. Acad. Sci. U. S. A.*, **109**,
40 13908–13914, doi:10.1073/pnas.1210106109. <http://www.ncbi.nlm.nih.gov/pubmed/22912403>
41 (Accessed October 29, 2018).
- 42 —, E. F. Lambin, and A. Reenberg, 2007: The emergence of land change science for global
43 environmental change and sustainability. *Proc. Natl. Acad. Sci. U. S. A.*, **104**, 20666–20671,
44 doi:10.1073/pnas.0704119104. <http://www.ncbi.nlm.nih.gov/pubmed/18093934> (Accessed
45 October 3, 2018).
- 46 Turner, B. L. (Billie L., W. C. Clark, R. W. Kates, J. F. Richards, T. Mathews, Jessica, and W. B.
47 Meyer, eds., 1990: *The Earth as transformed by human action : global and regional changes in
48 the biosphere over the past 300 years*. Cambridge University Press with Clark University,
49 Cambridge, UK and New York, USA, 713 pp.

- 1 Turner, K. G., and Coauthors, 2016: A review of methods, data, and models to assess changes in the
2 value of ecosystem services from land degradation and restoration. *Ecol. Modell.*, **319**, 190–207,
3 doi:10.1016/J.ECOLMODEL.2015.07.017.
4 <https://www.sciencedirect.com/science/article/pii/S0304380015003233> (Accessed February 21,
5 2019).
- 6 Turner, W., 2014: Conservation. Sensing biodiversity. *Science*, **346**, 301–302,
7 doi:10.1126/science.1256014. <http://www.ncbi.nlm.nih.gov/pubmed/25324372> (Accessed May
8 23, 2018).
- 9 Tzoulas, K., K. Korpela, S. Venn, V. Yli-Pelkonen, A. Kazmierczak, J. Niemela, and P. James, 2007:
10 Promoting ecosystem and human health in urban areas using Green Infrastructure: A literature
11 review. *Landsc. Urban Plan.*, **81**, 167–178, doi:10.1016/J.LANDURBPLAN.2007.02.001.
12 <https://www.sciencedirect.com/science/article/pii/S0169204607000503> (Accessed May 17,
13 2018).
- 14 Uddameri, V., S. Singaraju, and E. A. Hernandez, 2014: Impacts of sea-level rise and urbanization on
15 groundwater availability and sustainability of coastal communities in semi-arid South Texas.
16 *Environ. Earth Sci.*, **71**, 2503–2515, doi:10.1007/s12665-013-2904-z.
17 <http://link.springer.com/10.1007/s12665-013-2904-z> (Accessed March 29, 2019).
- 18 Uitto, J. I., 2016: Evaluating the environment as a global public good. *Evaluation*, **22**, 108–115,
19 doi:10.1177/1356389015623135. <http://journals.sagepub.com/doi/10.1177/1356389015623135>
20 (Accessed November 2, 2018).
- 21 Umunay, P. M., T. G. Gregoire, T. Gopalakrishna, P. W. Ellis, and F. E. Putz, 2019: Selective logging
22 emissions and potential emission reductions from reduced-impact logging in the Congo Basin.
23 *For. Ecol. Manage.*, **437**, 360–371, doi:10.1016/j.foreco.2019.01.049.
- 24 UNCCD, 1994: United Nations Convention to Combat Desertification.
- 25 —, 2016a: *Report of the Conference of the Parties on its twelfth session, held in Ankara from 12 to*
26 *23 October 2015.* Bonn, Germany,
27 https://www2.unccd.int/sites/default/files/sessions/documents/ICCD_COP12_20_Add.1/20add1
28 *eng.pdf.*
- 29 —, 2016b: *Land Degradation Neutrality Target Setting – A Technical Guide. Land degradation*
30 *neutrality target setting programme.* Bonn, Germany,
31 <https://www.unccd.int/sites/default/files/inline-files/LDN> TS Technical
32 *Guide_Draft_English.pdf.*
- 33 —, 2017: *The Global Land Outlook.* 1st ed. United Nations Convention to Combat
34 Desertification, Bonn, Germany, 340 pp.
- 35 UNDP, 2017: *Valuation of reforestation in terms of disaster risk reduction: A technical study from the*
36 *Republic of Korea.*
- 37 UNEP, 2007: *Sudan Post-Conflict Environmental Assessment.* United Nations Environment
38 Programme, Nairobi, Kenya, 358 pp. <http://www.unep.org> (Accessed May 15, 2018).
- 39 UNFCCC, 2013: Reporting and accounting of LULUCF activities under the Kyoto Protocol.
- 40 United Nations, 2015: *World Urbanization Prospects: The 2014 Revision. Department of Economic*
41 *and Social Affairs, New York.* https://esa.un.org/unpd/wup/publications/files/wup_pp.
- 42 Upadhyay, H., D. Mohan, and D. Mohan, 2017: Migrating to adapt? *Climate Change, Vulnerability*
43 *and Migration,* Routledge India, 43–58
44 <https://www.taylorfrancis.com/books/9781351375580/chapters/10.4324/978135147741-3>
45 (Accessed March 30, 2019).
- 46 Valade, A., V. Bellassen, C. Magand, and S. Luysaert, 2017: Sustaining the sequestration efficiency
47 of the European forest sector. *For. Ecol. Manage.*, **405**, 44–55,
48 doi:10.1016/j.foreco.2017.09.009.

- 1 VandenBygaart, A. J., 2016: The myth that no-till can mitigate global climate change. *Agric. Ecosyst.*
2 *Environ.*, **216**, 98–99, doi:10.1016/J.AGEE.2015.09.013.
3 <https://www.sciencedirect.com/science/article/pii/S0167880915300797> (Accessed May 8, 2018).
- 4 Vanek, S. J., and J. Lehmann, 2015: Phosphorus availability to beans via interactions between
5 mycorrhizas and biochar. *Plant Soil*, doi:10.1007/s11104-014-2246-y.
- 6 Vasconcellos, R. L. F., J. A. Bonfim, D. Baretta, and E. J. B. N. Cardoso, 2016: Arbuscular
7 Mycorrhizal Fungi and Glomalin-Related Soil Protein as Potential Indicators of Soil Quality in a
8 Recuperation Gradient of the Atlantic Forest in Brazil. *L. Degrad. Dev.*, **27**, 325–334,
9 doi:10.1002/ldr.2228. <http://doi.wiley.com/10.1002/ldr.2228> (Accessed March 16, 2019).
- 10 Vaughan, D. G., and Coauthors, 2013: Observations: Cryosphere. *Climate Change 2013: The*
11 *Physical Science Basis. Contribution of Working Group I to the Fifth Assessment Report of the*
12 *Intergovernmental Panel on Climate Change*, T.F. Stocker et al., Eds., Cambridge, UK and New
13 York, USA, p. 317-.
- 14 Vautard, R., J. Cattiaux, P. Yiou, J.-N. Thépaut, and P. Ciais, 2010: Northern Hemisphere
15 atmospheric stilling partly attributed to an increase in surface roughness. *Nat. Geosci.*, **3**, 756–
16 761, doi:10.1038/ngeo979. <http://www.nature.com/articles/ngeo979> (Accessed March 8, 2019).
- 17 Vavrus, S. J., F. He, J. E. Kutzbach, W. F. Ruddiman, and P. C. Tzedakis, 2018: Glacial Inception in
18 Marine Isotope Stage 19: An Orbital Analog for a Natural Holocene Climate. *Sci. Rep.*, **8**,
19 10213, doi:10.1038/s41598-018-28419-5. <http://www.nature.com/articles/s41598-018-28419-5>
20 (Accessed October 2, 2018).
- 21 Vecchi, G. A., K. L. Swanson, B. J. Soden, J. J. Sirutis, G. A. Vecchi, S. T. Garner, and I. M. Held,
22 2008: Climate change. Whither hurricane activity? *Science*, **322**, 687–689,
23 doi:10.1126/science.1164396. <http://www.ncbi.nlm.nih.gov/pubmed/18974337> (Accessed May
24 24, 2018).
- 25 Vedeld, P., A. Angelsen, J. Bojö, E. Sjaastad, and G. Kobugabe Berg, 2007: Forest environmental
26 incomes and the rural poor. *For. Policy Econ.*, **9**, 869–879,
27 doi:10.1016/J.FORPOL.2006.05.008.
28 <https://www.sciencedirect.com/science/article/pii/S1389934106001146> (Accessed March 8,
29 2019).
- 30 van der Ven, H., and B. Cashore, 2018: Forest certification: the challenge of measuring impacts. *Curr.*
31 *Opin. Environ. Sustain.*, **32**, 104–111, doi:10.1016/J.COSUST.2018.06.001.
32 <https://www.sciencedirect.com/science/article/pii/S1877343517302658> (Accessed March 29,
33 2019).
- 34 ———, C. Rothacker, and B. Cashore, 2018: Do eco-labels prevent deforestation? Lessons from non-
35 state market driven governance in the soy, palm oil, and cocoa sectors. *Glob. Environ. Chang.*,
36 **52**, 141–151, doi:10.1016/J.GLOENVCHA.2018.07.002.
37 <https://www.sciencedirect.com/science/article/abs/pii/S0959378017304545> (Accessed March
38 29, 2019).
- 39 Venter, O., and Coauthors, 2016: Sixteen years of change in the global terrestrial human footprint and
40 implications for biodiversity conservation. *Nat. Commun.*, **7**, 12558, doi:10.1038/ncomms12558.
41 <http://www.nature.com/doi/10.1038/ncomms12558> (Accessed October 25, 2018).
- 42 Ventura, M., and Coauthors, 2015: Biochar mineralization and priming effect on SOM decomposition
43 in two European short rotation coppices. *GCB Bioenergy*, doi:10.1111/gcbb.12219.
- 44 Verdone, M., and A. Seidl, 2017: Time, space, place, and the Bonn Challenge global forest restoration
45 target. *Restor. Ecol.*, **25**, 903–911, doi:10.1111/rec.12512.
46 <http://doi.wiley.com/10.1111/rec.12512> (Accessed April 11, 2019).
- 47 Verhoeven, E., E. Pereira, C. Decock, E. Suddick, T. Angst, and J. Six, 2017: Toward a Better
48 Assessment of Biochar–Nitrous Oxide Mitigation Potential at the Field Scale. *J. Environ. Qual.*,
49 doi:10.2134/jeq2016.10.0396.

- 1 Viger, M., R. D. Hancock, F. Miglietta, and G. Taylor, 2015: More plant growth but less plant
2 defence? First global gene expression data for plants grown in soil amended with biochar. *GCB*
3 *Bioenergy*, doi:10.1111/gcbb.12182.
- 4 Vincent, K. E., P. Tschakert, J. Barnett, M. G. Rivera-Ferre, and A. Woodward, 2014: Cross-chapter
5 box on gender and climate chan. *Climate Change 2014: Impacts, Adaptation, and Vulnerability.*
6 *Part A: Global and Sectoral Aspects. Contribution of Working Group II to the Fifth Assessment*
7 *Report of the Intergovernmental Panel on Climate Chang*, C.B. Field et al., Eds., Cambridge
8 University Press, Cambridge, UK and New York, NY, USA, 105–107.
- 9 Virapongse, A., B. A. Endress, M. P. Gilmore, C. Horn, and C. Romulo, 2017: Ecology, livelihoods,
10 and management of the *Mauritia flexuosa* palm in South America. *Glob. Ecol. Conserv.*, **10**, 70–
11 92, doi:10.1016/j.gecco.2016.12.005.
- 12 Viscarra Rossel, R. A., R. Webster, E. N. Bui, and J. A. Baldock, 2014: Baseline map of organic
13 carbon in Australian soil to support national carbon accounting and monitoring under climate
14 change. *Glob. Chang. Biol.*, **20**, 2953–2970, doi:10.1111/gcb.12569.
15 <http://doi.wiley.com/10.1111/gcb.12569> (Accessed February 27, 2019).
- 16 Vitousek, S., P. L. Barnard, C. H. Fletcher, N. Frazer, L. Erikson, and C. D. Storlazzi, 2017: Doubling
17 of coastal flooding frequency within decades due to sea-level rise. *Sci. Rep.*, **7**, 1399,
18 doi:10.1038/s41598-017-01362-7. <http://www.nature.com/articles/s41598-017-01362-7>
19 (Accessed March 18, 2019).
- 20 Vogl, A. L., B. P. Bryant, J. E. Hunink, S. Wolny, C. Apse, and P. Droogers, 2017: Valuing
21 investments in sustainable land management in the Upper Tana River basin, Kenya. *J. Environ.*
22 *Manage.*, **195**, 78–91, doi:10.1016/J.JENVMAN.2016.10.013.
23 <https://www.sciencedirect.com/science/article/pii/S030147971630785X> (Accessed April 17,
24 2019).
- 25 Vogt, J. V., U. Safriel, G. Von Maltitz, Y. Sokona, R. Zougmore, G. Bastin, and J. Hill, 2011:
26 Monitoring and assessment of land degradation and desertification: Towards new conceptual
27 and integrated approaches. *L. Degrad. Dev.*, **22**, 150–165, doi:10.1002/ldr.1075.
28 <http://doi.wiley.com/10.1002/ldr.1075> (Accessed September 26, 2018).
- 29 Volkova, L., H. Bi, J. Hilton, and C. J. Weston, 2017: Impact of mechanical thinning on forest carbon,
30 fuel hazard and simulated fire behaviour in *Eucalyptus delegatensis* forest of south-eastern
31 Australia. *For. Ecol. Manage.*, **405**, 92–100, doi:10.1016/j.foreco.2017.09.032.
- 32 —, S. H. Roxburgh, C. J. Weston, R. G. Benyon, A. L. Sullivan, and P. J. Polglase, 2018:
33 Importance of disturbance history on net primary productivity in the world’s most productive
34 forests and implications for the global carbon cycle. *Glob. Chang. Biol.*, **24**, 4293–4303,
35 doi:10.1111/gcb.14309.
- 36 van Wagner, C. E., 1978: Age-class distribution and the forest fire cycle. *Can. J. For. Res.*, **8**, 220–
37 227, doi:10.1139/x78-034. <http://www.nrcresearchpress.com/doi/10.1139/x78-034> (Accessed
38 May 26, 2018).
- 39 Wagner, L. E., 2013: A history of Wind Erosion Prediction Models in the United States Department
40 of Agriculture: The Wind Erosion Prediction System (WEPS). *Aeolian Res.*, **10**, 9–24,
41 doi:10.1016/J.AEOLIA.2012.10.001.
42 <https://www.sciencedirect.com/science/article/pii/S1875963712000389> (Accessed October 24,
43 2018).
- 44 Wairiu, M., 2017: Land degradation and sustainable land management practices in Pacific Island
45 Countries. *Reg. Environ. Chang.*, **17**, 1053–1064, doi:10.1007/s10113-016-1041-0.
- 46 Waldron, A., D. Garrity, Y. Malhi, C. Girardin, D. C. Miller, and N. Seddon, 2017: Agroforestry Can
47 Enhance Food Security While Meeting Other Sustainable Development Goals. *Trop. Conserv.*
48 *Sci.*, **10**, 194008291772066, doi:10.1177/1940082917720667.
49 <http://journals.sagepub.com/doi/10.1177/1940082917720667> (Accessed May 21, 2018).

- 1 Walsh, J. R., S. R. Carpenter, and M. J. Vander Zanden, 2016a: Invasive species triggers a massive
2 loss of ecosystem services through a trophic cascade. *Proc. Natl. Acad. Sci.*, **113**, 4081–4085,
3 doi:10.1073/pnas.1600366113. www.pnas.org/cgi/doi/10.1073/pnas.1600366113 (Accessed
4 October 1, 2018).
- 5 Walsh, K. J. E., and Coauthors, 2016b: Tropical cyclones and climate change. *Wiley Interdiscip. Rev.*
6 *Clim. Chang.*, **7**, 65–89, doi:10.1002/wcc.371. <http://doi.wiley.com/10.1002/wcc.371> (Accessed
7 March 17, 2019).
- 8 Walter Anthony, K., R. Daanen, P. Anthony, T. Schneider von Deimling, C.-L. Ping, J. P. Chanton,
9 and G. Grosse, 2016: Methane emissions proportional to permafrost carbon thawed in Arctic
10 lakes since the 1950s. *Nat. Geosci.*, **9**, 679–682, doi:10.1038/ngeo2795.
11 <http://www.nature.com/articles/ngeo2795> (Accessed March 16, 2019).
- 12 Wang, G., S. Mang, H. Cai, S. Liu, Z. Zhang, L. Wang, and J. L. Innes, 2016a: Integrated watershed
13 management: evolution, development and emerging trends. *J. For. Res.*, **27**, 967–994,
14 doi:10.1007/s11676-016-0293-3. <http://link.springer.com/10.1007/s11676-016-0293-3>
15 (Accessed April 17, 2019).
- 16 Wang, J., Z. Xiong, and Y. Kuzyakov, 2016b: Biochar stability in soil: meta-analysis of
17 decomposition and priming effects. *GCB Bioenergy*, **8**, 512–523, doi:10.1111/gcbb.12266.
18 <http://doi.wiley.com/10.1111/gcbb.12266> (Accessed December 1, 2017).
- 19 Wang, J., S. Yi, M. Li, L. Wang, and C. Song, 2018: Effects of sea level rise, land subsidence,
20 bathymetric change and typhoon tracks on storm flooding in the coastal areas of Shanghai. *Sci.*
21 *Total Environ.*, **621**, 228–234, doi:10.1016/J.SCITOTENV.2017.11.224.
22 <https://www.sciencedirect.com/science/article/pii/S0048969717332825> (Accessed March 17,
23 2019).
- 24 Wang, W., J. Sardans, C. Wang, C. Zeng, C. Tong, D. Asensio, and J. Peñuelas, 2017a: Relationships
25 between the potential production of the greenhouse gases CO₂, CH₄ and N₂O and soil
26 concentrations of C, N and P across 26 paddy fields in southeastern China. *Atmos. Environ.*,
27 **164**, 458–467, doi:10.1016/J.ATMOSENV.2017.06.023.
28 <https://www.sciencedirect.com/science/article/pii/S1352231017304041> (Accessed November 2,
29 2018).
- 30 Wang, Z., T. Hoffmann, J. Six, J. O. Kaplan, G. Govers, S. Doetterl, and K. Van Oost, 2017b:
31 Human-induced erosion has offset one-third of carbon emissions from land cover change. *Nat.*
32 *Clim. Chang.*, **7**, 345–349, doi:10.1038/nclimate3263.
33 <http://www.nature.com/doi/10.1038/nclimate3263> (Accessed March 16, 2019).
- 34 Ward, C., L. Stringer, and G. Holmes, 2018: Changing governance, changing inequalities: Protected
35 area co-management and access to forest ecosystem services: a Madagascar case study. *Ecosyst.*
36 *Serv.*, **30**, 137–148, doi:10.1016/J.ECOSER.2018.01.014.
37 <https://www.sciencedirect.com/science/article/pii/S2212041617304291> (Accessed October 28,
38 2018).
- 39 Ward, D., 2005: Do we understand the causes of bush encroachment in African savannas? *African J.*
40 *Range Forage Sci.*, **22**, 101–105, doi:10.2989/10220110509485867.
41 <http://www.tandfonline.com/doi/abs/10.2989/10220110509485867> (Accessed May 24, 2018).
- 42 Warren, A., 2002: Land degradation is contextual. *L. Degrad. Dev.*, **13**, 449–459, doi:10.1002/ldr.532.
43 <http://doi.wiley.com/10.1002/ldr.532> (Accessed May 16, 2018).
- 44 Warren, R., J. Price, J. VanDerWal, S. Cornelius, and H. Sohl, 2018: The implications of the United
45 Nations Paris Agreement on climate change for globally significant biodiversity areas. *Clim.*
46 *Change*, **147**, 395–409, doi:10.1007/s10584-018-2158-6.
47 <http://link.springer.com/10.1007/s10584-018-2158-6> (Accessed March 14, 2019).
- 48 Watson, R. T., I. R. Noble, B. Bolin, N. H. Ravindranath, D. J. Verardo, and D. J. Dokken, 2000:
49 *Land use, land-use change, and forestry*. R.T. Watson, I.R. Noble, B. Bolin, N.H. Ravindranath,

- 1 D.J. Verardo, and D.J. Dokken, Eds. Cambridge University Press, Cambridge, UK, 370 pp.
2 [http://98.131.92.124/sites/default/files/2000 Watson IPCC.pdf](http://98.131.92.124/sites/default/files/2000%20Watson%20IPCC.pdf) (Accessed May 16, 2018).
- 3 Webb, N. P., and Coauthors, 2016: The National Wind Erosion Research Network: Building a
4 standardized long-term data resource for aeolian research, modeling and land management.
5 *Aeolian Res.*, **22**, 23–36, doi:10.1016/J.AEOLIA.2016.05.005.
6 <https://www.sciencedirect.com/science/article/pii/S1875963716300568> (Accessed March 18,
7 2019).
- 8 —, and Coauthors, 2017a: Enhancing Wind Erosion Monitoring and Assessment for U.S.
9 Rangelands. *Rangelands*, **39**, 85–96, doi:10.1016/J.RALA.2017.04.001.
10 <https://www.sciencedirect.com/science/article/pii/S0190052817300160> (Accessed March 18,
11 2019).
- 12 —, N. A. Marshall, L. C. Stringer, M. S. Reed, A. Chappell, and J. E. Herrick, 2017b: Land
13 degradation and climate change: building climate resilience in agriculture. *Front. Ecol. Environ.*,
14 **15**, 450–459, doi:10.1002/fee.1530. <http://doi.wiley.com/10.1002/fee.1530> (Accessed December
15 20, 2017).
- 16 Wei, W., and Coauthors, 2016: Global synthesis of the classifications, distributions, benefits and
17 issues of terracing. *Earth-Science Rev.*, **159**, 388–403, doi:10.1016/J.EARSCIREV.2016.06.010.
18 <https://www.sciencedirect.com/science/article/pii/S0012825216301313> (Accessed October 28,
19 2018).
- 20 Wei, X., M. Shao, W. Gale, and L. Li, 2015: Global pattern of soil carbon losses due to the
21 conversion of forests to agricultural land. *Sci. Rep.*, **4**, 4062, doi:10.1038/srep04062.
22 <http://www.nature.com/articles/srep04062> (Accessed February 22, 2019).
- 23 Weichselgartner, J., and I. Kelman, 2015: Geographies of resilience. *Prog. Hum. Geogr.*, **39**, 249–
24 267, doi:10.1177/0309132513518834.
25 <http://journals.sagepub.com/doi/10.1177/0309132513518834> (Accessed March 30, 2019).
- 26 Weinzierl, T., J. Wehberg, J. Böhner, and O. Conrad, 2016: Spatial Assessment of Land Degradation
27 Risk for the Okavango River Catchment, Southern Africa. *L. Degrad. Dev.*, **27**, 281–294,
28 doi:10.1002/ldr.2426.
- 29 Weng, Z., and Coauthors, 2017: Biochar built soil carbon over a decade by stabilizing rhizodeposits.
30 *Nat. Clim. Chang.*, **7**, 371–376,
31 doi:10.1038/nclimate3276 <http://www.nature.com/nclimate/journal/v7/n5/abs/nclimate3276.html>
32 #supplementary-information. <http://dx.doi.org/10.1038/nclimate3276>.
- 33 Weng, Z. (Han), L. Van Zwieten, B. P. Singh, E. Tavakkoli, S. Kimber, S. Morris, L. M. Macdonald,
34 and A. Cowie, 2018: The accumulation of rhizodeposits in organo-mineral fractions promoted
35 biochar-induced negative priming of native soil organic carbon in Ferralsol. *Soil Biol. Biochem.*,
36 **118**, 91–96, doi:10.1016/j.soilbio.2017.12.008.
- 37 Weng, Z. H. (Han), L. Van Zwieten, B. P. Singh, S. Kimber, S. Morris, A. Cowie, and L. M.
38 Macdonald, 2015: Plant-biochar interactions drive the negative priming of soil organic carbon in
39 an annual ryegrass field system. *Soil Biol. Biochem.*, **90**, 111–121,
40 doi:10.1016/j.soilbio.2015.08.005.
41 <https://www.sciencedirect.com/science/article/pii/S0038071715002746> (Accessed December 2,
42 2017).
- 43 Wentworth, J., 2017: *Urban Green Infrastructure and Ecosystem Services*. London, UK, UK, 24 pp.
44 <https://researchbriefings.parliament.uk/ResearchBriefing/Summary/POST-PB-0026>.
- 45 van der Werf, G. R., and Coauthors, 2017: Global fire emissions estimates during 1997–2016. *Earth*
46 *Syst. Sci. Data*, **9**, 697–720, doi:10.5194/essd-9-697-2017.
- 47 Werner, F., R. Taverna, P. Hofer, E. Thürig, and E. Kaufmann, 2010: National and global greenhouse
48 gas dynamics of different forest management and wood use scenarios: a model-based
49 assessment. *Environ. Sci. Policy*, **13**, 72–85, doi:10.1016/j.envsci.2009.10.004.

- 1 <https://www.sciencedirect.com/science/article/pii/S1462901109001622> (Accessed May 26,
2 2018).
- 3 Wessels, K. ., S. . Prince, P. . Frost, and D. van Zyl, 2004: Assessing the effects of human-induced
4 land degradation in the former homelands of northern South Africa with a 1 km AVHRR NDVI
5 time-series. *Remote Sens. Environ.*, **91**, 47–67, doi:10.1016/j.rse.2004.02.005.
6 <https://www.sciencedirect.com/science/article/pii/S0034425704000653> (Accessed May 25,
7 2018).
- 8 Wessels, K. J., S. D. Prince, J. Malherbe, J. Small, P. E. Frost, and D. VanZyl, 2007: Can human-
9 induced land degradation be distinguished from the effects of rainfall variability? A case study
10 in South Africa. *J. Arid Environ.*, **68**, 271–297, doi:10.1016/j.jaridenv.2006.05.015.
11 <http://www.sciencedirect.com/science/article/pii/S014019630600190X> (Accessed October 19,
12 2017).
- 13 —, F. van den Bergh, and R. J. Scholes, 2012: Limits to detectability of land degradation by trend
14 analysis of vegetation index data. *Remote Sens. Environ.*, **125**, 10–22,
15 doi:10.1016/j.rse.2012.06.022.
16 <https://www.sciencedirect.com/science/article/pii/S0034425712002581> (Accessed May 25,
17 2018).
- 18 Westra, S., and Coauthors, 2014: Future changes to the intensity and frequency of short-duration
19 extreme rainfall. *Rev. Geophys.*, **52**, 522–555, doi:10.1002/2014RG000464.
20 <http://doi.wiley.com/10.1002/2014RG000464> (Accessed October 26, 2018).
- 21 White, J. W., G. Hoogenboom, B. A. Kimball, and G. W. Wall, 2011: Methodologies for simulating
22 impacts of climate change on crop production. *F. Crop. Res.*, **124**, 357–368,
23 doi:10.1016/J.FCR.2011.07.001.
24 <https://www.sciencedirect.com/science/article/pii/S0378429011002395> (Accessed May 31,
25 2018).
- 26 Wicke, B., E. Smeets, V. Dornburg, B. Vashev, T. Gaiser, W. Turkenburg, and A. Faaij, 2011: The
27 global technical and economic potential of bioenergy from salt-affected soils. *Energy Environ.*
28 *Sci.*, **4**, 2669, doi:10.1039/c1ee01029h.
- 29 —, P. Verweij, H. van Meijl, D. P. van Vuuren, and A. P. Faaij, 2012: Indirect land use change:
30 review of existing models and strategies for mitigation. *Biofuels*, **3**, 87–100,
31 doi:10.4155/bfs.11.154. <https://www.tandfonline.com/doi/full/10.4155/bfs.11.154> (Accessed
32 October 22, 2018).
- 33 Widgren, M., and J. E. G. Sutton, 2004: *Islands of intensive agriculture in Eastern Africa : past*
34 *& present*. Ohio University Press, Currey, Oxford, 160 pp. [http://www.diva-](http://www.diva-portal.org/smash/record.jsf?pid=diva2%3A185221&dswid=-9114)
35 [portal.org/smash/record.jsf?pid=diva2%3A185221&dswid=-9114](http://www.diva-portal.org/smash/record.jsf?pid=diva2%3A185221&dswid=-9114) (Accessed March 27, 2019).
- 36 Wiczorek, A. J., 2018: Sustainability transitions in developing countries: Major insights and their
37 implications for research and policy. *Environ. Sci. Policy*, **84**, 204–216,
38 doi:10.1016/J.ENVSCI.2017.08.008.
39 <https://www.sciencedirect.com/science/article/pii/S1462901117308092> (Accessed October 31,
40 2018).
- 41 Wiesmair, M., A. Otte, and R. Waldhardt, 2017: Relationships between plant diversity, vegetation
42 cover, and site conditions: implications for grassland conservation in the Greater Caucasus.
43 *Biodivers. Conserv.*, **26**, 273–291, doi:10.1007/s10531-016-1240-5.
44 <http://link.springer.com/10.1007/s10531-016-1240-5> (Accessed October 1, 2018).
- 45 Wigley, B. J., W. J. Bond, and M. T. Hoffman, 2010: Thicket expansion in a South African savanna
46 under divergent land use: local vs. global drivers? *Glob. Chang. Biol.*, **16**, 964–976,
47 doi:10.1111/j.1365-2486.2009.02030.x. <http://doi.wiley.com/10.1111/j.1365-2486.2009.02030.x>
48 (Accessed February 16, 2018).
- 49 Van Wilgen, B. W., N. Govender, H. C. Biggs, D. Ntsala, and X. N. Funda, 2004: Response of

- 1 Savanna Fire Regimes to Changing Fire-Management Policies in a Large African National Park.
2 *Conserv. Biol.*, **18**, 1533–1540, doi:10.1111/j.1523-1739.2004.00362.x.
3 <http://doi.wiley.com/10.1111/j.1523-1739.2004.00362.x> (Accessed May 8, 2018).
- 4 Wilhelm, J. A., and R. G. Smith, 2018: Ecosystem services and land sparing potential of urban and
5 peri-urban agriculture: A review. *Renew. Agric. Food Syst.*, **33**, 481–494,
6 doi:10.1017/S1742170517000205.
7 https://www.cambridge.org/core/product/identifier/S1742170517000205/type/journal_article
8 (Accessed October 31, 2018).
- 9 Wilson, D., and Coauthors, 2016: Greenhouse gas emission factors associated with rewetting of
10 organic soils. *Mires Peat*, doi:10.19189/MaP.2016.OMB.222.
- 11 Wilson, G. A., and Coauthors, 2017: Social Memory and the Resilience of Communities Affected by
12 Land Degradation. *L. Degrad. Dev.*, **28**, 383–400, doi:10.1002/ldr.2669.
13 <http://doi.wiley.com/10.1002/ldr.2669> (Accessed April 6, 2019).
- 14 Winder, R., E. Nelson, and T. Beardmore, 2011: Ecological implications for assisted migration in
15 Canadian forests. *For. Chron.*, **87**, 731–744, doi:10.5558/tfc2011-090.
16 <https://doi.org/10.5558/tfc2011-090>.
- 17 WOCAT, WOCAT (World overview of conservation approaches and technologies). *Glossary.*,
18 <https://www.wocat.net/en/glossary>.
- 19 Wong, P. P., I. J. Losada, J. P. Gattuso, J. Hinkel, A. Khattabi, K. L. McInnes, Y. Saito, and A.
20 Sallenger, 2014: Coastal systems and low-lying areas. *Climate Change 2014: Impacts,*
21 *Adaptation, and Vulnerability*, C.B. Field et al., Eds., Cambridge University Press, Cambridge
22 and New York, 361–409.
- 23 Woods, J., L. R. Lynd, M. Laser, M. Batistella, V. D. de Castro, K. Kline, and A. Faaij, 2015: Land
24 and Bioenergy. *Bioenergy & Sustainability: bridging the gaps - Scope Bioenergy*, G.M. Souza,
25 R.L. Victoria, C.A. Joly, and L.M. Verdade, Eds., SCOPE, Paris, 258–300
26 [http://www.bioenfapesp.org/scopebioenergy/index.php/32-uncategorised/100-scope-synthesis-](http://www.bioenfapesp.org/scopebioenergy/index.php/32-uncategorised/100-scope-synthesis-volume)
27 [volume](http://www.bioenfapesp.org/scopebioenergy/index.php/32-uncategorised/100-scope-synthesis-volume) (Accessed April 12, 2019).
- 28 Woolf, D., J. E. Amonette, F. A. Street-Perrott, J. Lehmann, and S. Joseph, 2010: Sustainable biochar
29 to mitigate global climate change. *Nat. Commun.*, **1**, 56.
- 30 —, and Coauthors, 2018: Biochar for Climate Change Mitigation. *Soil and Climate*, CRC Press,
31 Boca Raton, FL : CRC Press, Taylor & Francis Group, 2018. | Series: Advances in soil science,
32 219–248 <https://www.taylorfrancis.com/books/9780429945458/chapters/10.1201/b21225-8>
33 (Accessed April 7, 2019).
- 34 World Bank, 2009: *Environmental crisis or sustainable development opportunity? Transforming the*
35 *charcoal sector in Tanzania*. Washington, DC,
36 [https://openknowledge.worldbank.org/bitstream/handle/10986/2865/551400ESW0P1201PE1Ch](https://openknowledge.worldbank.org/bitstream/handle/10986/2865/551400ESW0P1201PE1Charcoal1TZ1FINAL.pdf?sequence=1)
37 [arcoal1TZ1FINAL.pdf?sequence=1](https://openknowledge.worldbank.org/bitstream/handle/10986/2865/551400ESW0P1201PE1Charcoal1TZ1FINAL.pdf?sequence=1).
- 38 —, 2016: *The Role of Green Infrastructure Solutions in Urban Flood Risk Management*.
39 Washington DC, USA, USA, 18 pp.
40 [http://documents.worldbank.org/curated/en/841391474028584058/text/108291-WP-P156654-](http://documents.worldbank.org/curated/en/841391474028584058/text/108291-WP-P156654-PUBLIC-ABSTRACT-SENT-WBUFCOPKnowledgeNoteGreenInfrastructureSolutions.txt)
41 [PUBLIC-ABSTRACT-SENT-WBUFCOPKnowledgeNoteGreenInfrastructureSolutions.txt](http://documents.worldbank.org/curated/en/841391474028584058/text/108291-WP-P156654-PUBLIC-ABSTRACT-SENT-WBUFCOPKnowledgeNoteGreenInfrastructureSolutions.txt).
- 42 Wu, H., and Coauthors, 2017a: The interactions of composting and biochar and their implications for
43 soil amendment and pollution remediation: a review. *Crit. Rev. Biotechnol.*, **37**, 754–764,
44 doi:10.1080/07388551.2016.1232696. <https://doi.org/10.1080/07388551.2016.1232696>.
- 45 Wu, J., J. Xiong, C. Hu, Y. Shi, K. Wang, and D. Zhang, 2015: Temperature sensitivity of soil
46 bacterial community along contrasting warming gradient. *Appl. Soil Ecol.*, **94**, 40–48,
47 doi:10.1016/J.APSSOIL.2015.04.018.
48 <https://www.sciencedirect.com/science/article/pii/S0929139315300044> (Accessed May 2, 2018).

- 1 Wu, Y., E. Chan, J. R. Melton, and D. L. Versegny, 2017b: A map of global peatland distribution
2 created using machine learning for use in terrestrial ecosystem and earth system models. *Geosci.*
3 *Model Dev. Discuss*, doi:10.5194/gmd-2017-152.
- 4 Wu, Z., P. Dijkstra, G. W. Koch, J. Peñuelas, and B. A. Hungate, 2011: Responses of terrestrial
5 ecosystems to temperature and precipitation change: a meta-analysis of experimental
6 manipulation. *Glob. Chang. Biol.*, **17**, 927–942, doi:10.1111/j.1365-2486.2010.02302.x.
7 <http://doi.wiley.com/10.1111/j.1365-2486.2010.02302.x> (Accessed May 23, 2018).
- 8 Wuest, S. B., J. D. Williams, and H. T. Gollany, 2006: Journal of soil and water conservation. *J. Soil*
9 *Water Conserv.*, **61**, 218–223. <http://www.jswnonline.org/content/61/4/218.short> (Accessed
10 October 18, 2018).
- 11 Xiao, L., X. Yang, H. Cai, L. Xiao, X. Yang, and H. Cai, 2017: The Indirect Roles of Roads in Soil
12 Erosion Evolution in Jiangxi Province, China: A Large Scale Perspective. *Sustainability*, **9**, 129,
13 doi:10.3390/su9010129. <http://www.mdpi.com/2071-1050/9/1/129> (Accessed March 5, 2019).
- 14 Xie, J., and Coauthors, 2016: Ten-year variability in ecosystem water use efficiency in an oak-
15 dominated temperate forest under a warming climate. *Agric. For. Meteorol.*, **218–219**, 209–217,
16 doi:10.1016/J.AGRFORMET.2015.12.059.
17 <https://www.sciencedirect.com/science/article/pii/S0168192315300204> (Accessed March 14,
18 2019).
- 19 Xu, G., Y. Zhang, H. Shao, and J. Sun, 2016: Pyrolysis temperature affects phosphorus transformation
20 in biochar: Chemical fractionation and ³¹P NMR analysis. *Sci. Total Environ.*,
21 doi:10.1016/j.scitotenv.2016.06.081.
- 22 Xu, J., R. Yin, Z. Li, and C. Liu, 2006: China's ecological rehabilitation: Unprecedented efforts,
23 dramatic impacts, and requisite policies. *Ecol. Econ.*, doi:10.1016/j.ecolecon.2005.05.008.
- 24 Xu, J., P. J. Morris, J. Liu, and J. Holden, 2018a: PEATMAP: Refining estimates of global peatland
25 distribution based on a meta-analysis. *Catena*, **160**, 134–140, doi:10.1016/j.catena.2017.09.010.
- 26 Xu, X., K. Cheng, H. Wu, J. Sun, Q. Yue, and G. Pan, 2019: Greenhouse gas mitigation potential in
27 crop production with biochar soil amendment—a carbon footprint assessment for cross-site field
28 experiments from China. *GCB Bioenergy*, **11**, 592–605, doi:10.1111/gcbb.12561.
29 <http://doi.wiley.com/10.1111/gcbb.12561> (Accessed April 7, 2019).
- 30 Xu, Z., C. E. Smyth, T. C. Lemprière, G. J. Rampley, and W. A. Kurz, 2018b: Climate change
31 mitigation strategies in the forest sector: biophysical impacts and economic implications in
32 British Columbia, Canada. *Mitig. Adapt. Strateg. Glob. Chang.*, **23**, 257–290,
33 doi:10.1007/s11027-016-9735-7. <http://link.springer.com/10.1007/s11027-016-9735-7>
34 (Accessed May 26, 2018).
- 35 Yamanoi, K., Y. Mizoguchi, and H. Utsugi, 2015: Effects of a windthrow disturbance on the carbon
36 balance of a broadleaf deciduous forest in Hokkaido, Japan. *Biogeosciences*, **12**, 6837–6851,
37 doi:10.5194/bg-12-6837-2015.
- 38 Yang, J., G. Cao, D. Han, H. Yuan, Y. Hu, P. Shi, and Y. Chen, 2019: Deformation of the aquifer
39 system under groundwater level fluctuations and its implication for land subsidence control in
40 the Tianjin coastal region. *Environ. Monit. Assess.*, **191**, 162, doi:10.1007/s10661-019-7296-4.
41 <http://link.springer.com/10.1007/s10661-019-7296-4> (Accessed March 17, 2019).
- 42 Yang, M., F. E. Nelson, N. I. Shiklomanov, D. Guo, and G. Wan, 2010: Permafrost degradation and
43 its environmental effects on the Tibetan Plateau: A review of recent research. *Earth-Science*
44 *Rev.*, **103**, 31–44, doi:10.1016/J.EARSCIREV.2010.07.002.
45 <https://www.sciencedirect.com/science/article/pii/S0012825210000826> (Accessed May 23,
46 2018).
- 47 Yang, Y., R. J. Donohue, T. R. McVicar, M. L. Roderick, and H. E. Beck, 2016: Long-term CO₂
48 fertilization increases vegetation productivity and has little effect on hydrological partitioning in
49 tropical rainforests. *J. Geophys. Res. Biogeosciences*, **121**, 2125–2140,

- 1 doi:10.1002/2016JG003475. <http://doi.wiley.com/10.1002/2016JG003475> (Accessed October
2 24, 2018).
- 3 Yao, L., L. Chen, W. Wei, and R. Sun, 2015: Potential reduction in urban runoff by green spaces in
4 Beijing: A scenario analysis. *Urban For. Urban Green.*, doi:10.1016/j.ufug.2015.02.014.
- 5 Yengoh, G. T., and J. Ardö, 2014: Crop Yield Gaps in Cameroon. *Ambio*, **43**, 175–190,
6 doi:10.1007/s13280-013-0428-0. <http://link.springer.com/10.1007/s13280-013-0428-0>
7 (Accessed October 31, 2018).
- 8 —, D. Dent, L. Olsson, A. Tengberg, and C. J. Tucker, 2015: *Use of the normalized difference*
9 *vegetation index (NDVI) to assess land degradation at multiple scales : current status, future*
10 *trends, and practical considerations*. Springer, Heidelberg, New York, Dordrecht, London, 110
11 pp.
- 12 Yin, R., 2009: *An Integrated Assessment of China's Ecological Restoration Programs*. Springer,
13 New York.
- 14 Young, A., 1995: *Agroforestry for soil conservation*. CTA, Wageningen, The Netherlands, 194 pp.
15 <https://www.cabdirect.org/cabdirect/abstract/19986773686> (Accessed October 22, 2018).
- 16 Young, A., 1999: Is there Really Spare Land? A Critique of Estimates of Available Cultivable Land in
17 Developing Countries. *Environ. Dev. Sustain.*, **1**, 3–18, doi:10.1023/A:1010055012699.
18 <http://link.springer.com/10.1023/A:1010055012699> (Accessed April 11, 2019).
- 19 Young, E., D. Muir, A. Dawson, and S. Dawson, 2014: Community driven coastal management: An
20 example of the implementation of a coastal defence bund on South Uist, Scottish Outer
21 Hebrides. *Ocean Coast. Manag.*, **94**, 30–37, doi:10.1016/j.ocecoaman.2014.01.001.
- 22 Yu, H., and Coauthors, 2015: The fertilizing role of African dust in the Amazon rainforest: A first
23 multiyear assessment based on data from Cloud-Aerosol Lidar and Infrared Pathfinder Satellite
24 Observations. *Geophys. Res. Lett.*, **42**, 1984–1991, doi:10.1002/2015GL063040.
25 <http://doi.wiley.com/10.1002/2015GL063040> (Accessed February 27, 2019).
- 26 Zabel, F., B. Putzenlechner, and W. Mauser, 2014: Global Agricultural Land Resources – A High
27 Resolution Suitability Evaluation and Its Perspectives until 2100 under Climate Change
28 Conditions. *PLoS One*, **9**, e107522, doi:10.1371/journal.pone.0107522.
29 <http://dx.plos.org/10.1371/journal.pone.0107522> (Accessed May 22, 2018).
- 30 Zahawi, R. A., G. Duran, and U. Kormann, 2015: Sixty-Seven Years of Land-Use Change in Southern
31 Costa Rica. *PLoS One*, **10**, e0143554, doi:10.1371/journal.pone.0143554.
32 <http://dx.plos.org/10.1371/journal.pone.0143554> (Accessed October 1, 2018).
- 33 Zedler, J. B., 2000: Progress in wetland restoration ecology. *Trends Ecol. Evol.*, **15**, 402–407,
34 doi:10.1016/S0169-5347(00)01959-5.
35 <https://www.sciencedirect.com/science/article/pii/S0169534700019595> (Accessed May 24,
36 2018).
- 37 Zeng, Z., and Coauthors, 2017: Climate mitigation from vegetation biophysical feedbacks during the
38 past three decades. *Nat. Clim. Chang.*, **7**, 432–436, doi:10.1038/nclimate3299.
39 <http://www.nature.com/articles/nclimate3299> (Accessed March 16, 2019).
- 40 Zhang, M., and Coauthors, 2017: A global review on hydrological responses to forest change across
41 multiple spatial scales: Importance of scale, climate, forest type and hydrological regime. *J.*
42 *Hydrol.*, **546**, 44–59, doi:10.1016/J.JHYDROL.2016.12.040.
43 <https://www.sciencedirect.com/science/article/pii/S0022169416308307> (Accessed October 27,
44 2018).
- 45 Zhang, P., G. Shao, G. Zhao, D. C. Le Master, G. R. Parker, J. B. Dunning, and Q. Li, 2000: China's
46 forest policy for the 21st century. *Science (80-.)*, doi:10.1126/science.288.5474.2135.
- 47 Zhang, X., and Coauthors, 2013: Using biochar for remediation of soils contaminated with heavy
48 metals and organic pollutants. *Environ. Sci. Pollut. Res.*, doi:10.1007/s11356-013-1659-0.

- 1 Zhang, X. C., and M. A. Nearing, 2005: Impact of climate change on soil erosion, runoff, and wheat
2 productivity in central Oklahoma. *CATENA*, **61**, 185–195, doi:10.1016/J.CATENA.2005.03.009.
3 <https://www.sciencedirect.com/science/article/pii/S0341816205000494> (Accessed May 23,
4 2018).
- 5 Zhou, Z., C. Wang, and Y. Luo, 2018: Effects of forest degradation on microbial communities and
6 soil carbon cycling: A global meta-analysis. *Glob. Ecol. Biogeogr.*, **27**, 110–124,
7 doi:10.1111/geb.12663. <http://doi.wiley.com/10.1111/geb.12663> (Accessed September 27,
8 2018).
- 9 Zhu, Q., X. Yang, B. Yu, M. Tulau, S. McInnes-Clarke, R. H. Nolan, Z. Du, and Q. Yu, 2019:
10 Estimation of event-based rainfall erosivity from radar after wildfire. *L. Degrad. Dev.*, **30**, 33–
11 48, doi:10.1002/ldr.3146. <http://doi.wiley.com/10.1002/ldr.3146> (Accessed March 8, 2019).
- 12 Zimmerer, K. S., 1993: Soil Erosion and Social (Dis)Courses in Cochabamba, Bolivia: Perceiving the
13 Nature of Environmental Degradation. *Econ. Geogr.*, **69**, 312, doi:10.2307/143453.
14 <https://www.jstor.org/stable/143453?origin=crossref> (Accessed April 28, 2018).
- 15 Zölch, T., J. Maderspacher, C. Wamsler, and S. Pauleit, 2016: Using green infrastructure for urban
16 climate-proofing: An evaluation of heat mitigation measures at the micro-scale. *Urban For.*
17 *Urban Green.*, **20**, 305–316, doi:10.1016/J.UFUG.2016.09.011.
18 <https://www.sciencedirect.com/science/article/pii/S1618866716301686> (Accessed May 17,
19 2018).
- 20 Zomer, R. J., H. Neufeldt, J. Xu, A. Ahrends, D. Bossio, A. Trabucco, M. van Noordwijk, and M.
21 Wang, 2016: Global Tree Cover and Biomass Carbon on Agricultural Land: The contribution of
22 agroforestry to global and national carbon budgets. *Sci. Rep.*, **6**, 29987, doi:10.1038/srep29987.
23 <http://www.nature.com/articles/srep29987> (Accessed April 3, 2019).
- 24 Zulu, L. C., 2010: The forbidden fuel: Charcoal, urban woodfuel demand and supply dynamics,
25 community forest management and woodfuel policy in Malawi. *Energy Policy*, **38**, 3717–3730,
26 doi:<https://doi.org/10.1016/j.enpol.2010.02.050>.
27 <http://www.sciencedirect.com/science/article/pii/S0301421510001540>.
- 28 Zulu, L. C., and R. B. Richardson, 2013: Charcoal, livelihoods, and poverty reduction: Evidence from
29 sub-Saharan Africa. *Energy Sustain. Dev.*, **17**, 127–137,
30 doi:<https://doi.org/10.1016/j.esd.2012.07.007>.
31 <http://www.sciencedirect.com/science/article/pii/S0973082612000506>.
- 32 Van Zwieten, L., S. Kimber, S. Morris, K. Y. Chan, A. Downie, J. Rust, S. Joseph, and A. Cowie,
33 2010: Effects of biochar from slow pyrolysis of papermill waste on agronomic performance and
34 soil fertility. *Plant Soil*, **327**, 235–246, doi:10.1007/s11104-009-0050-x.
35 <http://link.springer.com/10.1007/s11104-009-0050-x> (Accessed April 2, 2019).
- 36 Van Zwieten, L., T. Rose, D. Herridge, S. Kimber, J. Rust, A. Cowie, and S. Morris, 2015: Enhanced
37 biological N₂ fixation and yield of faba bean (*Vicia faba* L.) in an acid soil following biochar
38 addition: dissection of causal mechanisms. *Plant Soil*, **395**, 7–20, doi:10.1007/s11104-015-2427-
39 3. <http://link.springer.com/10.1007/s11104-015-2427-3> (Accessed April 2, 2019).
- 40